Empirical evidence of non-linearity in bottom up effect in a marine predator-prey system

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Abstract

Strength of species interaction may have profound effects on population dynamics. Empirical estimates of interaction strength is often based on the assumption that the interaction strengths are constant. Barents Sea cod and capelin are two fish populations for which such interaction has been acknowledged and used, under the assumption of constant interaction strength, when studying their population dynamics. However, species interaction can often be non-linear in marine ecosystems and might profoundly change our understanding of food chains. Analysing 37 years long survey time series in the Arcto-Boreal Barents Sea with a state-space modelling framework, we demonstrate that the effect of capelin on cod is not linear but shifts depending on capelin abundance: while capelin is beneficial for cod population at high abundance, below the threshold, it becomes less important for cod. Our analysis therefore shows the importance of investigating non-linearity in species interaction and may contribute to an improved understanding on species assemblages.

**Keywords**: Bayesian statistic, Gompertz model, species competition, predation, threshold modelling

1. Introduction

Climate change is profoundly affecting and altering marine systems (1). Indirect effects of climate change, such as alteration of species interactions, might have a stronger impact on population dynamics than the direct warming effects (2, 3). Environment can also have a non-additive effect (e.g., threshold) on population dynamics in terrestrial (4, 5) and marine (6-10) systems alike resulting in different population equilibrium and dynamics (4). Marine systems are prone to non-linear transitions under climate warming (1) and overfishing (11) that may also lead to altered population dynamics (12, 13). A prime example of such non-linear transition is the Atlantic cod (10, 11, 14). Such non-linear transition was however seldom studied in relation to species interactions (but see 12). To study such interactions, Wootton and Emmerson (15) suggest to use long-term time series, take into account non linearity and process errors. This can be achieved using state-space modelling approaches (12, 16) in data rich systems such as the Barents Sea (17).

Here we explore the population dynamics of two interacting species: Barents Sea (BS) capelin *Mallotus villosus* and Northeast Arctic (NEA) cod *Gadus morhua*. Both species are known to interact in the Barents Sea and affect each other’s population (18). Indeed, predation by NEA cod on BS capelin is thought to have delayed the capelin stock’s recovery after its collapses (13). In addition, the BS capelin is considered to be the main food for NEA cod (19, 20) and low capelin stock was blamed for the very low cod catches at the end of 1980s (21). Both species population dynamics are well documented to be affected by environmental variables (e.g. 22, 23).

Here, we applied a Gompertz state-space model (12, 16) on 37-years long survey time series of BS capelin and NEA cod (24) aiming at (i) assessing whether there is a linear or non-linear interaction between cod and capelin, (ii) understanding what a non-linear dynamics means for the population and the trophic interactions in the system.

2. Methods

We analysed jointly the change in population abundance for the Barents Sea (BS) capelin and Northeast Arctic (NEA) cod from the Barents Sea (figure 1). Population data (1981-2019) were published fish stock assessments data (table 9.4 for the capelin, table A3 for the cod) (24). Capelin stock size in numbers are estimates from the August-September acoustic survey, and cod abundance are indices in numbers from the January-March bottom trawl surveys in the Barents Sea (figure 2).

In addition, we used two climatic variables (the Kola transect sea temperature, ST, and the winter North Atlantic Oscillation, wNAO) as potential environmental drivers of capelin and cod population dynamics (e.g. 22, 23). The sea temperature (1921-2019) is an aggregated average over the upper 200 m at five stations (3-7) on the Kola meridian transect (33°30'E, 70°30'-72°30'N) in the Barents Sea (http://www.pinro.ru/; 25). The December to March North Atlantic Oscillation index (26) represents North Atlantic-scale climate effects (1964-2019). Both ST and NAO were standardized to z-scores.

**Model description**

The analyses were based on a Gompertz state-space model (12) reparametrized as in Stenseth et al. (27) incorporating competition (intra- and interspecific respectively ai,i (with the intra-specific interaction set to 1 (27, 28)) and ai,j) and environmental variables (ai,st and ai,nao) effects. The model (table 1: eqn. 1) incorporated also a Gaussian distributed stochastic term (ε) to acknowledge our inadequate understanding of the population *i* dynamicscomplexity(i.e., the process error: eqn. 2).

Since the sampling of capelin population is in August-September while in January-March for the cod, the cod survey at yearyr was conducted between the capelin surveys at yearyr-1 and yearyr. We took this into account when modelling the capelin by using the cod abundance estimate at yearyr (Nj,yr) instead of yearyr-1 (Nj,yr-1) as described in eqn.3.

We assumed that the observed abundances (Obs; from trawl survey for cod and acoustic survey for capelin) were normally distributed (in log scale) with variance term σ2i,obsaround the true log population values for the species *i* (eqn. 4). Prior specifications are found in the accompanying codes in the supplementary material.

To detect possible non-linear dynamics, we tested for potential pairwise interactions between all explanatory variables (table 1) using the Bürmann’s expansion (29). In short, the Bürmann’s expansion test checks interaction between pairs of variables by analysing the residuals between additive models with or without interaction thus finding the best fit and reports significance. Only when non-linearity was detected did we include a threshold non-additive effect in the Gompertz state-space model(12). In our case (see results) the threshold non-additive effect let the growth potential of species *i* and the effect of species *j* on species *i* (ai,0 /*bj0* and ai,j */bij* respectively) change according to whether the threshold variable (X) was below or above some threshold level θ (eqn. 5).

To detect if and at what value the covariate X has a meaningful threshold effect, the model calculates the log-likelihood of the process equation (i.e. the underlying population dynamics) for each value of X in the data (i.e., capelin abundance ln(Ncap), see codes in supplementary materials). A threshold is identified when a single value θ of X produces a large spike in log-likelihood (figure S3). In which case, the threshold value is located somewhere between ≥θb and <θ with θb being the first value lower than the selected θ. Moreover, to remove any “border” effects i.e. spurious detection of a threshold value due to very unequal partitioning of the data (e.g. 95% below threshold *vs.* 5% above), our model searched for a potential threshold value only within the 20-80 percentiles of the available values of X (21 values used out of 37).

We used a Bayesian Markov Chain Monte Carlo (MCMC) approach to jointly estimate all parameters (for both capelin and cod) in a single model for the period 1981-2019. We used the *Stan* *software* via the R packages *rstan* (v.2.21.3) and *shinystan* (30). A likelihood function was created based on the model and data, and in combination with the prior distributions of the parameters, the posterior distributions were estimated. Weakly informative priors were used in order to let the data drive the inferences except for the process and observation error variances. The latter were not identifiable alone thus we included an informative prior on the ratio of the process to observation error variance centred around 1 (*Normal*(1, 0.5)) (31). A sensitivity test with a ratio centred around 0.5 and 2 (respectively *Normal*(0.5, 0.5) and *Normal*(2, 0.5)) showed that the choice of the exact value did not affect our results (figure S1). Note that there were no indication of correlation between the estimated process errors of the two species and hence they were modelled as such (figure S2).

We used four independent chains with 50,000 iterations each, where the first 30,000 iterations were discarded as ”burn-in” iterations to ensure that the chains had converged. In addition, we thinned the chains with a factor 10 to reduce autocorrelation in the posterior samples and to produce a reasonable amount of output. We used the Gelman and Rubin Ȓ convergence diagnostics (32) and visual inspection of the chains to ensure convergence, and posterior predictive checks to evaluate the model fit. All analyses were conducted using the software R v.4.1.3 (33).

3. Results

The Bürmann-test indicated an interaction between capelin and cod abundance for both capelin and cod models (p < 0.05). We first used non-additive models to describe the dynamics of both species (eqn. 5) but only the model for cod showed a relevant threshold (figure S3). We then modelled capelin following eqn. 3 and cod following eqn. 5 (table 1). The cod model estimated a threshold θ between > 201 109 and ≤ 209 109 capelins (figure S3).

Model convergence was evaluated by visual inspection of the four parallel Hamiltonian Monte Carlo chains. The chains were well mixed, had low autocorrelation after thinning and displayed no trends after the burn-in iterations. There were no divergent transitions in the chains. The Gelman and Rubin Ȓ convergence diagnostics were <1.002 for all model parameters thus supporting convergence. In addition, there was no systematic deviation between the fitted values and the observed time series (figure 2).

Median parameter estimates from these models are presented in table 2 (see figure S4 for the full marginal posterior distributions). As expected, the model indicated a positive effect of the previous year abundance for both species. The environmental variables (ST and wNAO) did not show an effect for cod but ST showed an effect (positive) for capelin. For capelin, the cod showed a negative effect indicating a predation effect.

For cod, the capelin abundance, as expressed in number of capelin, has a biologically important effect. The interspecific competition term – the effect of capelin numbers on cod – was negligible (acod,cap) when capelin was under the capelin stock size threshold and was positive (bcod,cap) over the threshold changing from 0.03 to 0.52 in a log scale (table 2 and figure S4).

This indicated that the capelin abundance had an effect on cod population only when the capelin stock was big enough (over 209 billionindividuals).

4. Discussion

Through the use of a state-space model, that combined long-term population time series with environmental variables, we illustrated how historically established species interactions may be drastically modified if explored for nonlinearity. . Especially, we find empirical evidence for nonlinear change in species interaction (table 2) directly linked to prey abundance change. Non-additive population dynamics has been previously described for many species, notably for cod due to this species data availability (11, 34) but seldom addressing interaction with another species (12, 13).

The NEA cod is a predator of the BS capelin as shown by diet studies (19, 20) and we indeed found a negative effect of cod on capelin stock, similar to previous findings (35). Conversely, capelin abundance is expected to have a positive effect on cod stock (36, 37) and our results also support the claim. However, they also indicate that capelin effect on cod is non-linear and it becomes negligible for low capelin abundance.

Capelin is highly represented in the cod diet during warm years, with temperature affecting both BS capelin’s distribution (38) and recruitment (35). However, cod is a generalist predator with a diet following the community composition change (39). Indeed, the composition of the cod diet changes over time in response to environmental conditions and the dynamics of the prey populations (20, 40). This is particularly visible for the capelin proportion in the cod diet that follows the capelin population change, hence its availability as prey for the cod. This high plasticity in its diet may explain our result that cod population is not affected by capelin abundance when the latter is under a relatively high threshold of 209 billion individuals (note that the median capelin abundance during the studied period is 227 billion individuals, data ranging from 14 billion to 1016 billion individuals). In addition, a low capelin abundance has been associated to high herring *Clupea harengus* abundance, another major predator of capelin larvae in the Barents Sea (35, 41) that is also part of the cod diet (19, 20).

Our model takes into account the main processes affecting the dynamics of a population i.e. inter-specific competition, intra-specific competition (i.e., density dependence), and environmental conditions. However, our model does not take into account the spatial overlap of the two species that affects their interaction (42) neither the effect of the potential interaction with other species of the system (e.g. haddock Melanogrammus aeglefinus 12, herring Clupea harengus 13, Polar cod Boreogadus saida 43). These lacking processes are however partially taken into account by the process error in the model formulation (15) (see figure S5).

In this study, we show that a nonlinearity in the species interactions has an impact on population dynamics and affects our understanding of the functioning of the food chain similar to what was observed for the effect of climate warming (6, 12). Stock assessment are conducted on a single species basis but incorporates more and more some known interaction between the species of interest and climate or other species (44). For instance, BS capelin is managed by taking into account the NEA cod predation (24). Given the implication our results can have on the understanding of NEA cod population dynamics, our approach could be timely and necessary.

Data accessibility. Data are freely available in the report of the Arctic Fisheries Working Group (AFWG) 2019 of the International Council for the Exploration of the Sea (24) at

https://ices-library.figshare.com/articles/report/Arctic\_Fisheries\_Working\_Group\_AFWG\_/18618752?file=33397001; wNAO at https://climatedataguide.ucar.edu/climate-data/hurrell-north-atlantic-oscillation-nao-index-station-based. Kola sea temperature from Polar branch of the Russian Federal Institute of Fisheries and Oceanography at http://www.pinro.ru/ (45).

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Competing interests. We have no competing interests.

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1. Hoegh-Guldberg O, Bruno JF. The Impact of Climate Change on the World's Marine Ecosystems. Science. 2010;328(5985):1523-8.

2. Ockendon N, Baker DJ, Carr JA, White EC, Almond REA, Amano T, et al. Mechanisms underpinning climatic impacts on natural populations: altered species interactions are more important than direct effects. Glob Change Biol. 2014;20(7):2221-9.

3. Stige LC, Ono K, Eriksen E, Dalpadado P. Direct and indirect effects of sea ice cover on major zooplankton groups and planktivorous fishes in the Barents Sea. ICES J Mar Sci. 2019;76(1):24-36.

4. Stenseth NC, Durant JM, Fowler MS, Matthysen E, Adriaensen F, Jonzén N, et al. Testing for effects of climate change on competitive relationships and coexistence between two bird species. Proc R Soc B. 2015;282(1807).

5. Stenseth NC, Chan KS, Tavecchia G, Coulson T, Mysterud A, Clutton-Brock T, et al. Modelling non-additive and nonlinear signals from climatic noise in ecological time series: Soay sheep as an example. Proceedings of the Royal Society B-Biological Sciences. 2004;271(1552):1985-93.

6. Ciannelli L, Fisher JAD, Skern-Mauritzen M, Hunsicker ME, Hidalgo M, Frank KT, et al. Theory, consequences and evidence of eroding population spatial structure in harvested marine fishes: a review. Mar Ecol Progr Ser. 2013;480:227-43.

7. Dingsør GE, Ciannelli L, Chan KS, Ottersen G, Stenseth NC. Density dependence and density independence during the early life stages of four marine fish stocks. Ecology. 2007;88(3):625-34.

8. Llope M, Daskalov GM, Rouyer TA, Mihneva V, Chan KS, Grishin AN, et al. Overfishing of top predators eroded the resilience of the Black Sea system regardless of the climate and anthropogenic conditions. Glob Change Biol. 2011;17(3):1251-65.

9. Lindegren M, Checkley Jr DM, Koslow JA, Goericke R, Ohman MD. Climate-mediated changes in marine ecosystem regulation during El Niño. Glob Change Biol. 2018;24(2):796-809.

10. Vasilakopoulos P, Marshall CT. Resilience and tipping points of an exploited fish population over six decades. Glob Change Biol. 2015;21(5):1834-47.

11. Sguotti C, Otto SA, Frelat R, Langbehn TJ, Plambech Ryberg M, Lindegren M, et al. Catastrophic dynamics limit Atlantic cod recovery. Proceedings of the Royal Society B: Biological Sciences. 2019;286(1898):20182877.

12. Durant JM, Ono K, Stenseth NC, Langangen Ø. Non-linearity in interspecific interactions in response to climate change: cod and haddock as an example. Glob Change Biol. 2020;26:5554– 63.

13. Durant JM, Aarvold L, Langangen Ø. Stock collapse and its effect on species interactions: Cod and herring in the Norwegian-Barents Seas system as an example. Ecology and Evolution. 2021;11(23):16993-7004.

14. Frank KT, Petrie B, Fisher JAD, Leggett WC. Transient dynamics of an altered large marine ecosystem. Nature. 2011;477:86-9.

15. Wootton JT, Emmerson M. Measurement of Interaction Strength in Nature. Annual Review of Ecology, Evolution, and Systematics. 2005;36:419-44.

16. Ono K, Langangen Ø, Stenseth NC. Improving risk assessments in conservation ecology. Nature Communications. 2019;10(1):2836.

17. Paasche O, Osterblom H, Neuenfeldt S, Bonsdorff E, Brander K, Conley DJ, et al. Connecting the Seas of Norden. Nature Climate Change. 2015;5(2):89-92.

18. Lilly GR, Nakken O, Brattey J. A review of the contributions of fisheries and climate variability to contrasting dynamics in two Arcto-boreal Atlantic cod (*Gadus morhua*) stocks: Persistent high productivity in the Barents Sea and collapse on the Newfoundland and Labrador Shelf. Prog Oceanogr. 2013;114:106-25.

19. Townhill BL, Holt RE, Bogstad B, Durant JM, Pinnegar JK, Dolgov AV, et al. Diets of the Barents Sea cod (*Gadus morhua*) from the 1930s to 2018. Earth Syst Sci Data. 2021;13(3):1361-70.

20. Holt RE, Bogstad B, Durant JM, Dolgov AV, Ottersen G. Barents Sea cod (*Gadus morhua*) diet composition: long-term interannual, seasonal, and ontogenetic patterns. ICES J Mar Sci. 2019;76(6):1641-52.

21. Hjermann DØ, Ottersen G, Stenseth NC. Competition among fishermen and fish causes the collapse of Barents Sea capelin. Proc Natl Acad Sci U S A. 2004;101(32):11679-84.

22. Hjermann DØ, Bogstad B, Eikeset AM, Ottersen G, Gjosaeter H, Stenseth NC. Food web dynamics affect Northeast Arctic cod recruitment. Proceedings of the Royal Society B-Biological Sciences. 2007;274(1610):661-9.

23. Durant JM, Hjermann DØ. Age-structure, harvesting and climate effects on population growth of Arcto-boreal fish stocks. Mar Ecol Prog Ser. 2017;577:177-88.

24. ICES. Arctic Fisheries Working Group (AFWG). ICES Scientific Reports. 2019;01(30):pp 934.

25. Ingvaldsen R, Loeng H, Ådlandsvik B, Ottersen G. Climate variability in the Barents Sea during the 20th century with focus on the 1990s. ICES Marine Science Symposia. 2003;219:160- 8.

26. Hurrell JW, Deser C. North Atlantic climate variability: The role of the North Atlantic Oscillation. J Mar Syst. 2009;78(1):28-41.

27. Stenseth NC, Durant JM, Fowler MS, Matthysen E, Adriaensen F, Jonzén N, et al. Testing for effects of climate change on competitive relationships and coexistence between two bird species. Proceedings of the Royal Society B-Biological Sciences. 2015;282(1807):20141958.

28. Mutshinda CM, O’Hara RB, Woiwod IP. A multispecies perspective on ecological impacts of climatic forcing. J Anim Ecol. 2011;80(1):101-7.

29. Chan KS, Kristoffersen AB, Stenseth NC. Bürmann expansion and test for additivity. Biometrika. 2003;90:209-22.

30. Stan Development Team. RStan: the R interface to Stan. R package version 2.21.3. 2021;https://mc-stan.org/.

31. Ono K, Punt AE, Rivot E. Model performance analysis for Bayesian biomass dynamics models using bias, precision and reliability metrics. Fisheries Research. 2012;125-126:173-83.

32. Gelman A, Rubin DB. Inference from iterative simulation using multiple sequences. Statistical Science. 1992;7:457-72.

33. R Core Team. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. 2022:URL http://www.R-project.org.

34. Fauchald P. Predator-prey reversal: A possible mechanism for ecosystem hysteresis in the North Sea? Ecology. 2010;91(8):2191-7.

35. Hjermann DØ, Bogstad B, Dingsor GE, Gjosaeter H, Ottersen G, Eikeset AM, et al. Trophic interactions affecting a key ecosystem component: a multistage analysis of the recruitment of the Barents Sea capelin (*Mallotus villosus*). Can J Fish Aquat Sci. 2010;67(9):1363-75.

36. Hjermann DØ, Stenseth NC, Ottersen G. The population dynamics of Northeast Arctic cod (*Gadus morhua*) through two decades: an analysis based on survey data. Can J Fish Aquat Sci. 2004;61(9):1747-55.

37. Durant JM, Hjermann DØ, Sabarros PS, Stenseth NC. Northeast arctic cod population persistence in the Lofoten-Barents Sea system under fishing. Ecol Appl. 2008;18(3):662-9.

38. Hop H, Gjøsæter H. Polar cod (*Boreogadus saida*) and capelin (*Mallotus villosus*) as key species in marine food webs of the Arctic and the Barents Sea. Mar Biol Res. 2013;9(9):878-94.

39. Ellingsen KE, Yoccoz NG, Tveraa T, Frank KT, Johannesen E, Anderson MJ, et al. The rise of a marine generalist predator and the fall of beta diversity. Glob Change Biol. 2020:doi: 10.1111/gcb.15027.

40. Durant JM, Skern-Mauritzen M, Krasnov YV, Nikolaeva NG, Lindstrom U, Dolgov A. Temporal dynamics of top predators interactions in the Barents Sea. PLoS ONE. 2014;9(11):e110933.

41. Hallfredsson EH, Pedersen T. Effects of predation from juvenile herring (*Clupea harengus*) on mortality rates of capelin (*Mallotus villosus*) larvae. Can J Fish Aquat Sci. 2009;66(10):1693-706.

42. Fall J, Ciannelli L, Skaret G, Johannesen E. Seasonal dynamics of spatial distributions and overlap between Northeast Arctic cod (*Gadus morhua*) and capelin (*Mallotus villosus*) in the Barents Sea. PLoS ONE. 2018;13(10):e0205921.

43. Dupont N, Durant JM, Gjøsæter H, Langangen Ø, Stige LC. Effects of sea ice cover, temperature and predation on the stock dynamics of the key Arctic fish species polar cod *Boreogadus saida*. Mar Ecol Prog Ser. 2021;677:141-59.

44. Skern-Mauritzen M, Ottersen G, Handegard NO, Huse G, Dingsor GE, Stenseth NC, et al. Ecosystem processes are rarely included in tactical fisheries management. Fish Fish. 2016;17(1):165-75.

45. Tereschenko VV. Seasonal and year-to-year variations of temperature and salinity along the Kola meridian transect. ICES CM. 1996;C11:pp24.

46. Bakketeig IE, Hauge M, Kvamme C, Sunnset BH, Toft KØ. Havforskningsrapporten 2016. Fisken Havet. 2016;1(2016):99.

**Table 1.** Summary of the equations.

|  |  |  |  |
| --- | --- | --- | --- |
| **Formulation** | | | **Eqn.** |
| ln(Ni,yr) = |  | ai,0 + ai,i·ln(Ni,yr-1) + ai,j·ln(Nj,yr-1)+ ai,st·STyr-1+ ai,nao·wNAOyr-1 + εi,yr | 1 |
| εi,yr ~ *Normal*(0, σ2i,proc) | | | 2 |
| ln(Ni,yr) = |  | ai,0 + ai,i·ln(Ni,yr-1) + ai,j·ln(Nj,yr)+ ai,st·STyr-1+ ai,nao·wNAOyr-1 + εi,yr | 3 |
| ln(Obsi,yr) ~ *Normal*(ln(Ni,yr), σ2i,obs) | | | 4 |
| ln(Ni,yr) = | { | ai,0 + ai,i·ln(Ni,yr-1) + ai,j·ln(Nj,yr-1)+ ai,st·STyr-1+ ai,nao·wNAOyr-1 + εi,yr, if ln(Nj,yr-1) < θ | 5 |
| bi,0 + ai,i·ln(Ni,yr-1) + bi,j·ln(Nj,yr-1)+ ai,st·STyr-1+ ai,nao·wNAOyr-1 + εi,yr, |

With Ni,yr the abundance for the species *i* at year *yr*. Process error εi,yr of the population *i* with variance σ2i,proc. Observation error ln(Obsi,yr) of the population *i* with variance σ2i,obs. ST=sea temperature, wNAO = winter North Atlantic Oscillation, a or b =parameterswith the first subscript the modelled species and the second subscript the variable – environment or other species – affecting this species.

**Table 2.** Estimated parameters for the two models.

|  |  |  |  |
| --- | --- | --- | --- |
| **Stock** | **Parameter** | **Median estimate** | **%** |
| **> 0** |
| NEA cod | acod,0 (cap < θ) | 0.94 | 71 |
| bcod,0 (cap ≥ θ) | -1.85 | 28 |
|  | acod,cod | 0.84 | 100 |
|  | acod,cap (cap < θ) | 0.03 | 55 |
|  | bcod,cap (cap ≥ θ) | 0.52 | 88 |
|  | acod,st | 0.06 | 71 |
|  | acod,nao | 0.05 | 81 |
| BS capelin | acap,0 | 5.74 | 100 |
|  | acap,cap | 0.56 | 100 |
|  | acap,cod | -0.49 | 2 |
|  | acap,st | 0.17 | 86 |
|  | acap,nao | -0.01 | 45 |

Subscripts “cap” stands for Barents Sea (BS) capelin and “cod” for Northeast Arctic (NEA) cod. θ is the threshold value (log-transformed BS capelin abundance of 5.34 109 individuals). Note that ca 50% indicates that the posterior values are centred around 0. See figure S4.

Figure captions

**Figure 1.** Approximate feeding distributions in the Barents Sea of the Northeast Arctic cod (blue) and the capelin (red). The map are redrawn from Bakketeig et al. 2016 (46).

**Figure 2.** Posterior predictive check on abundances indices. The dots show the acoustic (for capelin in A) and trawl (for cod in B) survey log-transformed abundance estimates (i.e. data) and the blue bands are the 95% predictive intervals. For both plots, filled data points are for the years with capelin stock size over the estimated threshold between > 5.30 109 (blue dotted line in figure A) and ≤ 5.34 109 (red dotted line in figure A).