## REPORT OF THE

## North-Western Working Group

## ICES Headquarters

29 April-8 May 2003

## PARTS 1 AND 2

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International Council for the Exploration of the Sea
Conseil International pour l'Exploration de la Mer

## TECHNICAL MINUTES

# North-Western Working Group (NWWG) 

## ACFM Meeting May 2003

Sub-group Chair: Dankert Skagen<br>Working Group Chair : Einar Hjörleifsson<br>Rapporteur/reviewer : Holger Hovgård<br>Reviewer : André Forest<br>Jesper Boje, Jakup Reinert, Ken Patterson

The WG chair introduced the report with a discussion on the various models used in the NWWG meeting in 2003. The WG found it considerable easier to inspect peculiarities of the data using some of the alternate approaches than by using the standard XSA software. The NWWG found that the standard software were not taken the development in assessment models and computer technology into account and included a number of stabilising techniques which final weight are based on ad hoc decisions that are not always transparent. The chair noted that some of new approaches (e.g. the AD-CAM) integrated the catch-at-age analysis, recruitment estimation, and short and medium-term forecasts all into one step and thus removing the inconsistencies and potential for error that arises when these are all separate steps. This approach also allows for stochastic forecasts and much more explicit statements about the uncertainty associated with a particular assessment.

The Sub-group supported the idea that an assessment should be more than following a standard template and pushing all the right buttons, but rather that an assessment should actually involve finding-out what the data tell you. The ICES approach can limit the extent to which this is possible, a point which will be raised in plenary. The Sub-group noted that the is a need for a description of the assumptions used in the various alternate model and there is a need for describing how to interpret input and output to facilitate the assessment reviews.

## General comments

The sub-group appreciated the fact that the WG use a broad range of assessment models. However, use of more flexible models requires that much more detail needs to be provided in terms of model specification and diagnostics than for a 'standard XSA'. This is often not the case for the models used here, making it very difficult to review some of the assessments.

The sub-group noted that the Yield-per-recruit analyses uses the long-term averages of weight, maturities and exploitation patterns. For weights and maturity this may be justified but it appears questionable for the exploitation rate.

In several cases important graphs are presented as very small. Considering the old age of the reviewers this may impede a thorough evaluation.

The NWWG often uses various models to predict mean weights in the projections. In several cases the models used are poorly described leaving little possibility for the reviewer to evaluate if the estimates are better than simple averages.

## Icelandic Stocks

The sub-group appreciated that a section on discard are provided for the Icelandic stocks. The estimated discard levels reported appears low considering the incentives for high grading in the ITQ regulation applied. Discards are not included in the assessment due to lack of historical information.

## Icelandic cod

The gillnet mesh size used should be given in cm , full mesh.

The working group appreciated the extensive use of simple descriptive statistics methods to scrutinize data and the various quality control plots comparing survey abundances and VPA estimates.

The work up of the survey indices is not clearly described. The wording "conventional Cochran type method" is ambiguous (should be read as a stratified random design).

The tuning fleets file used in the models should be explicitly tabled. For the XSA run it should be explained that the survey year is shifted except for the youngest age group.

Different measurements WEST are used for the SSB (first 5 month commercial catch data) and for the exploitable biomass $(\mathrm{B}+)$. The reason - that the $\mathrm{B}+$ measure relates to the established HCL - should have been given in the report. It should be considered to use the observed spring weights and maturities from the surveys instead of the commercial values.

The significant work associated with comparing a range of different models is appreciated and it is comforting that the main model provides similar results. The AD-CAM model chosen as a) statistically based and b) combining assessment and forecast.

There are blocks in the F's for the oldest ages for the AD-CAM model (T. 3.3.12). The reason for that should be explored.

## Icelandic Haddock

The reasoning in the text is sometimes difficult to follow and the text would benefit from a stringent editing. Table headings are in several cases poor. More digits are needed in the CAA table.

The report contains a full documentation of one XSA run but no tabled output from the chosen AD-CAM. This severely impedes the possibility for reviewing the quality of the assessment. The AD-CAM model allow a presentation of the CAA residuals that should be provided.

All examined model fails to provide good fits of the spring 2003 survey. The high survey biomass is expected caused by changes in haddock availability.

Given that two large surveys are available attempts should be given to use all the survey information available.

The sub group inspected the estimated stock numbers and mortalities from the final ADCAM run (tabled below). Apparently a very strong constraint has been applied on the year to year variation at the oldest ages. A justification for this should be given. Moreover, this constraint does not carry over to the prediction phase.

Table 1 Icelandic haddock. Estimates of population size at the beginning of the year. Estimates based on the ADCAM model (the adopted run used for the catch forecast)

| year/age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 46294 | 78691 | 119060 | 26335 | 19886 | 21236 | 3181 | 756 | 112 |
| 1980 | 11953 | 37902 | 64162 | 95220 | 18337 | 10910 | 8758 | 1149 | 244 |
| 1981 | 51577 | 9786 | 30750 | 51320 | 67929 | 10743 | 4808 | 3321 | 370 |
| 1982 | 37140 | 42228 | 7979 | 24610 | 36722 | 39768 | 4578 | 1810 | 1064 |
| 1983 | 23032 | 30408 | 34489 | 6339 | 17196 | 21261 | 18089 | 1754 | 575 |
| 1984 | 52210 | 18857 | 24839 | 27430 | 4138 | 9662 | 9051 | 7282 | 555 |
| 1985 | 105382 | 42746 | 15379 | 19479 | 17541 | 2193 | 3905 | 3709 | 2302 |
| 1986 | 206095 | 86280 | 34753 | 11397 | 11553 | 8648 | 838 | 1562 | 1157 |
| 1987 | 52232 | 168736 | 70327 | 25185 | 6211 | 5072 | 2963 | 321 | 483 |
| 1988 | 30275 | 42764 | 136790 | 51505 | 13883 | 2625 | 1854 | 1121 | 99 |
| 1989 | 28740 | 24787 | 34822 | 101966 | 29220 | 6074 | 944 | 686 | 344 |
| 1990 | 99943 | 23531 | 20161 | 26000 | 61021 | 14398 | 2143 | 341 | 209 |
| 1991 | 204106 | 81827 | 18937 | 14643 | 15166 | 29072 | 5529 | 766 | 104 |
| 1992 | 43360 | 167108 | 65228 | 14049 | 8428 | 6896 | 11199 | 1960 | 232 |
| 1993 | 47053 | 35500 | 134449 | 47151 | 7837 | 3584 | 2516 | 3834 | 592 |
| 1994 | 85293 | 38524 | 28788 | 98775 | 26689 | 3372 | 1315 | 831 | 1154 |
| 1995 | 43617 | 69832 | 31187 | 20742 | 57257 | 11575 | 1238 | 427 | 249 |
| 1996 | 117513 | 35710 | 55489 | 21127 | 11923 | 25983 | 4176 | 401 | 126 |
| 1997 | 17785 | 96211 | 28273 | 37555 | 11220 | 5635 | 9431 | 1334 | 118 |
| 1998 | 58095 | 14562 | 77323 | 19687 | 20251 | 4900 | 2223 | 3106 | 393 |
| 1999 | 144659 | 47564 | 11707 | 55291 | 10489 | 8734 | 1796 | 747 | 913 |
| 2000 | 173030 | 118437 | 38049 | 8202 | 30106 | 4215 | 3019 | 589 | 220 |
| 2001 | 190594 | 141665 | 94847 | 25745 | 4511 | 12755 | 1438 | 959 | 173 |
| 2002 | 49014 | 156045 | 114060 | 67673 | 14699 | 2251 | 4989 | 452 | 284 |
| 2003 | 138684 | 40129 | 126781 | 83629 | 40738 | 7355 | 868 | 1551 | 134 |
| 2004 | 63212 | 113545 | 32453 | 93711 | 52441 | 21912 | 3422 | 369 | 621 |
| 2005 | 63212 | 51754 | 91842 | 24024 | 58995 | 28382 | 10280 | 1468 | 149 |
| 2006 | 63212 | 51754 | 41862 | 67988 | 15124 | 31929 | 13315 | 4409 | 594 |
| 2007 | 63212 | 51754 | 41862 | 30989 | 42801 | 8185 | 14979 | 5711 | 1785 |

Table 2
Icelandic haddock. Estimates of fishing mortality. Estimates based on the ADCAM model (the adopted run used for the catch forecast)

| year/age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| F4-7 |  |  |  |  |  |  |  |  |  |
| 1979 | 0.00 | 0.02 | 0.16 | 0.40 | 0.69 | 0.82 | 0.93 | 0.93 | 0.516 |
| 1980 | 0.01 | 0.02 | 0.14 | 0.33 | 0.62 | 0.77 | 0.93 | 0.93 | 0.465 |
| 1981 | 0.00 | 0.02 | 0.13 | 0.34 | 0.65 | 0.78 | 0.94 | 0.94 | 0.475 |
| 1982 | 0.00 | 0.03 | 0.16 | 0.35 | 0.59 | 0.76 | 0.95 | 0.95 | 0.463 |
| 1983 | 0.00 | 0.03 | 0.23 | 0.38 | 0.65 | 0.71 | 0.95 | 0.95 | 0.492 |
| 1984 | 0.00 | 0.04 | 0.25 | 0.44 | 0.71 | 0.69 | 0.95 | 0.95 | 0.520 |
| 1985 | 0.01 | 0.10 | 0.32 | 0.51 | 0.76 | 0.72 | 0.97 | 0.97 | 0.577 |
| 1986 | 0.00 | 0.12 | 0.41 | 0.62 | 0.87 | 0.76 | 0.97 | 0.97 | 0.666 |
| 1987 | 0.01 | 0.11 | 0.40 | 0.66 | 0.81 | 0.77 | 0.98 | 0.98 | 0.659 |
| 1988 | 0.01 | 0.09 | 0.37 | 0.63 | 0.82 | 0.80 | 0.98 | 0.98 | 0.653 |
| 1989 | 0.01 | 0.09 | 0.31 | 0.51 | 0.84 | 0.82 | 0.99 | 0.99 | 0.620 |
| 1990 | 0.02 | 0.12 | 0.34 | 0.54 | 0.76 | 0.83 | 0.99 | 0.99 | 0.617 |
| 1991 | 0.03 | 0.10 | 0.35 | 0.59 | 0.75 | 0.84 | 0.99 | 0.99 | 0.633 |
| 1992 | 0.02 | 0.12 | 0.38 | 0.66 | 0.81 | 0.87 | 1.00 | 1.00 | 0.680 |
| 1993 | 0.01 | 0.11 | 0.37 | 0.64 | 0.80 | 0.91 | 1.00 | 1.00 | 0.681 |
| 1994 | 0.01 | 0.13 | 0.35 | 0.64 | 0.80 | 0.92 | 1.01 | 1.01 | 0.677 |
| 1995 | 0.03 | 0.19 | 0.35 | 0.59 | 0.82 | 0.93 | 1.02 | 1.02 | 0.673 |
| 1996 | 0.03 | 0.19 | 0.43 | 0.55 | 0.81 | 0.94 | 1.02 | 1.02 | 0.684 |
| 1997 | 0.02 | 0.16 | 0.42 | 0.63 | 0.73 | 0.91 | 1.02 | 1.02 | 0.672 |
| 1998 | 0.02 | 0.14 | 0.43 | 0.64 | 0.80 | 0.89 | 1.02 | 1.02 | 0.691 |
| 1999 | 0.02 | 0.16 | 0.41 | 0.71 | 0.86 | 0.92 | 1.02 | 1.02 | 0.724 |
| 2000 | 0.02 | 0.19 | 0.40 | 0.66 | 0.88 | 0.95 | 1.02 | 1.02 | 0.720 |
| 2001 | 0.02 | 0.14 | 0.36 | 0.50 | 0.74 | 0.96 | 1.01 | 1.01 | 0.638 |
| 2002 | 0.01 | 0.11 | 0.31 | 0.49 | 0.75 | 0.97 | 1.01 | 1.01 | 0.630 |
| 2003 | 0.01 | 0.10 | 0.27 | 0.42 | 0.57 | 0.66 | 0.72 | 0.72 | 0.477 |
| 2004 | 0.01 | 0.10 | 0.26 | 0.41 | 0.56 | 0.65 | 0.70 | 0.70 | 0.470 |
| 2005 | 0.01 | 0.10 | 0.26 | 0.41 | 0.56 | 0.65 | 0.70 | 0.70 | 0.470 |
| 2006 | 0.01 | 0.10 | 0.26 | 0.41 | 0.56 | 0.65 | 0.70 | 0.70 | 0.470 |
| 2007 | 0.01 | 0.10 | 0.26 | 0.41 | 0.56 | 0.65 | 0.70 | 0.70 | 0.470 |
|  |  |  |  |  |  |  |  |  |  |

## Farose Plateau Cod

Considering the little difference in the various individual tuning fleets the WG may consider conducting runs including all potential tuning fleets.

The high value of F this year is surprising as anecdotal information may indicate no drastic increase in fishing effort. Questionable how much confidence can be placed on the high F given that the retrospective indicates that F is consistently overestimated. The high variability in F may to some extent be expected considering the low value of shrinkage used. The review group conducted two exploratory runs a) a run with fixed q's on the older ages and b) a run applying a more heavy shrinkage of 0.5 . There were no differences between the NWWG run and the fixed q run whereas the 0.5 shrinkage run gave lower Fbar's for the last years. The WG is encouraged to further evaluate the effect of the XSA settings at its next meeting.

There are signs of changes in the residuals for age 5-6 over the survey time-series that should be investigated. Given that the only tuning fleets are one surveys potential changes in availabilities should be investigated. However, the reason for these trends may also be found in the catch data series that appears very noisy especially for the older ages.

The subgroup would welcome if more descriptive diagnostic were presented, e.g. plots of survey year class size versus VPA estimates.

## Faroese Bank cod

There is apparently no incentives for misreporting catches to the bank in recent years. This may place some confidence to the amount caught and reported from the Bank.

The use of the ratio of the catch to the survey CPUE hinges on differences between the two indices in the earlier period as the two indices is highly correlated for the reason years. It should be asserted that both the surveys and the catch figures indices are reliable for the up to 1990.

## Faroese Haddock

The shifting of the spring survey should have been explicitly stated in the text.

The spring survey has been made available since last year and included as requested from ACFM.

The WG should ascertain that the expansion of the age group down to the age 0 do not affect the survivor estimates of the older age groups. It could be considered to remove the youngest age group in each survey as the XSA residuals are considerable.

## Faroese Saithe

Again this year the final run is not provided. There is in fact little reason to run a hand tuned VPA. Revision could be restricted to the forecast.

There are trends in the residuals with large negative residuals in the earlier years. Considering that this a commercial fleet that are subjected to technological changes the WG group should consider to shorten the time-series used.

The setting of the 1998 YC is sat conservatively at 80 mill. Actual knowledge of the YC and basically based on the CAA. The survey estimates (summer) do however confirm that the YC is large.

## Greenland Cod

The sub-group appreciated the effort to assess the inshore stock components. Given that the fishery presently is limited there may be little scope for conducting assessments based on CAA (mortality assumed to be driven by natural mortality). Information on sampling levels should be provided.

Survey data may be appropriate for evaluating the age structure of the inshore stocks. However, the present gillnet survey are only efficient for catching younger cod. Addition of larger mesh sizes in the net series may be considered.

However, it may also be considered to increase the depth strata covered as the present survey is restricted to shallow water.

The Sub-Group appreciates the work conducted on defining reference points for the offshore stock. The suggested measures -rescaled B4+ biomass from a log-log regression appears somewhat in transparent.

## Greenland Halibut

The working may consider to investigate possible technological creeping in the fisheries as that may be of considerable importance when using commercial catch rates in the production model.

Catch expectations for 2003 are assumed as 33000 tons (revised from 30 Kt in WG report).

## Redfish

For information on how the splitting of the commercial catches on species/stocks readers are referred to the 1998 WG report. A short outline on the procedures will facilitate the review. Similarly, comments on possible problems and potential biases caused by the splitting procedures will be appreciated.

## Marinus

The sub-group appreciates the work carried out to use BORNICOM as a tool for assessing S. marinus. If this model is to be accepted for assessing the stock a proper review it requires more comprehensive presentations of settings, input and results.

It clearly appears that recruitment varies considerable and that the stock depends on a few outstanding year classes, i.e. 1985 and 1990. As abundant year classes are only discernable from ca. age 10 (Table 8.1.5) this imply that there are only 2 known poor YC's since the latest good one. The available data do therefore not suggest immediate recruitment crises.

## Deep sea mentella on the continental slope

The review group had few comments on this stock that was assessed using similar procedures as last year.

## Pelagic mentella

The stock status is obviously very difficult to evaluate due to its wide geographical distribution and the considerable number of fleets that exploit it. The review group also notes considerable uncertainties in the historical catch statistics.

The assessment is based on an international survey combining acoustics and trawling. No survey data were available for 2002. A comprehensive international survey is to be conducted in summer 2003 and there may therefore be reasons to revisit the assessment when this data become available.

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### 1.1 Participants

| Einar Hjörleifsson (Chair) | Iceland |
| :--- | :--- |
| Jesper Boje | Greenland |
| Höskuldur Björnsson | Iceland |
| Luis Rideao Crus | Faroe Islands |
| Fernando Gonzalez | Spain |
| Agnes C. Gundersen | Norway |
| Aage S. Høines | Norway (part-time) |
| Kristján Kristinsson | Iceland |
| Jean-Jacques Maguire | Faroe Islands |
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| Lise Helen Ofstad | Faroe Islands |
| Marie Storr-Paulsen | Greenland |
| Jákup Reinert | Faroe Islands |
| Porsteinn Sigurdsson | Iceland |
| Björn Ævarr Steinarsson | Iceland |
| Petur Steingrund | Faroe Islands |

### 1.2 Terms of Reference

The Northwestern Working Group [NWWG] (Chair: E. Hjorleifsson, Iceland) will meet at ICES Headquarters from 29 April to 8 May 2003 to:
a) assess the status of and provide catch options for 2004 for the stocks of redfish in Subareas V, XII and XIV; Greenland halibut in Subareas V and XIV; cod in Subarea XIV, NAFO Subarea 1, and Division Va; saithe in Division Va; and haddock in Division Va;
b) for cod, haddock and saithe in Division Vb , that are under effort control, assess the status of and provide effort options and expected corresponding catches for 2004;
c) update survey and fishery information on the stocks of redfish in Subareas V, VI, XII and XIV. In particular, update information on the development of the pelagic fishery for redfish with respect to seasonal and area distribution to allow NEAFC to further consider the appropriateness of area and seasonal closures;
d) consider further possibilities for the incorporation of biological interactions into the assessments of capelin, herring, and cod stocks in Division Va;
e) update information on the stock composition, distribution and migration of the redfish stocks in Subareas V and XIV, and comment on the possible relationship between pelagic "deep sea" Sebastes mentella and the Sebastes mentella fished in demersal fisheries on the continental shelf and slope;
f) provide information on the horizontal and vertical distribution of pelagic redfish stock components in the Irminger Sea as well as seasonal and interannual changes in distribution;
g) provide specific information on possible deficiencies in the assessments including at least: Major inadequacies in the data on catches, effort or discards; major inadequacies if any in research vessel surveys data and major difficulties if any in model formulation; including inadequacies in available software. The Group should clarify the consequences from these deficiencies for $a$ ) assessment of the status of the stocks and $b$ ) for the projection;
h) for stocks for which a full analytical assessment is presented, comment on this meeting's assessments compared to the last assessment of the same stock;
i) comment on the PA reference points proposed by the Study Group on Precautionary Reference Points for Advice on Fishery Management;
j) structure the assessment report following the guidelines as adopted by ACFM in October 2002 with special attention to the quality issues.

NWWG will report by 9 May 2003 for the attention of ACFM.

## Request from NEAFC to ICES for scientific advice for 2004

In addition to the ToR from ICES the NWWG addressed the NEAFC requests to ICES on the following issue:

1. Regarding redfish stocks:
a) Submit new information on stock identity of the components of redfish such as "pelagic deep-sea" Sebastes mentella, "oceanic" Sebastes mentella fished in the pelagic fisheries, and the "deep-sea" Sebastes mentella fished in demersal fisheries on the continental shelf and slope;
b) Provide information on the horizontal and vertical distribution of pelagic redfish and fisheries in the Irminger Sea and adjacent waters as well as seasonal and interannual changes in distribution;
c) Comment on whether the horizontal, vertical and seasonal distribution of pelagic redfish in the Irminger Sea indicates the presence of different stock components within the area.

### 1.3 Report structure

The format of the report is similar to the years 1999-2001, with Tables and Figures located after all text for each stock. In the 1999 report some information not used directly in the assessment was omitted in order to make it more digestible for clients. This year the text on the Sebastes mentella was reduced and hopefully clarified, this is in response to readers' request to make it more digestible. Other systematic attempts to reduce the amount of documentation have not been made in the last four years' reports.

Based on the limited experimentation by the Working Group it was concluded that the adopted ACFM guidelines are an improvement of the current practice (ToRj). During the meeting it became apparent that the change in the structure of the report could not be completed within the time frame of the meeting and thus intersessional work will be done to complete the task.

### 1.4 Stocks assessed by NWWG

### 1.4.1 Introduction

The stocks dealt with by NWWG can be divided into two classes: those for which data are sufficient to allow an agebased analytical assessment, and those for which either the data is limited or for which the quality of the data is questionable, impeding analytical assessments. All gadoid stocks are in the first class except for Faroe Bank cod, where a short time-series and incomplete biological sampling of the landings inhibit standard ICES analytical assessment, and the offshore cod in Greenland, where a ceased fishery prevents a VPA (At this meeting exploration of incomplete catch data from the inshore cod in Greenland were done for the first time by applying a statistical approach). In the second class are most of the redfish management units as well as Greenland halibut. One redfish stock, S. marinus, sits in the middle of these two extremes, being assessed by a length-based model (Bormicon).

### 1.4.2 Age-based analytical assessements

For most of the stocks for which age-based analytical assessments were carried out, the terminal fishing mortality was estimated by tuning aged catch data with selected fleet age-disaggregated commercial or survey indices. In the final run only the Faroe saithe was based on a commercial tuning series since no reliable survey index is available for that stock.

### 1.4.2.1 Faroe stocks

The assessment on the Faroese stocks has been based on the Lowestoft software. In last year's assessment the Faroe summer survey tuning fleet was available for the first time. In last year's assessment the point estimators carried forward into the predictions for cod and haddock were based on using only those indices and exclude all commercial indices. The longer Faroe spring survey series, which has been extensively reworked in the past years was made available to the WG for the first time this year. After some preliminary analysis it was concluded that it was justified to include the spring survey in the tuning of Faroe haddock. Due to large residuals in cod indices from the spring survey, especially in older age groups, it was decided not to use that survey in the tuning of that stock. At this Working Group meeting various alternatives to the XSA were explored in particular to i) study some of the peculiarities of the XSA outcome that may be related to the assumptions in the model, and ii) get a sense of the noise in the input data. Time as well as unfamiliarity of some of the members to the methods used did not permit us to include the various analyses in the Working Group report. However, the advantages of e.g. statistical catch-at-age compared with the XSA model were obvious, particularly in relation to establishing a more reasonable fishing pattern in recent years.

### 1.4.2.2 Icelandic stocks

The Icelandic saithe was not assessed at this Working Group meeting. In recent years both the cod and haddock have been assessed by using various software packages. The reason for the use of different software is a result of the preference and expertise of the individual user that does the assessment. All the models are based on catch-at-age analysis (i.e. using the stock and the catch equation) using survey information as additional information. Various different assumptions are then explored by the different individuals running the different software - the final choice of settings by each person is based on personal judgement (sometimes referred to as expert opinion). The point estimators from the different models are thus not driven by the name of the model but by the assumptions behind each model!

The results from the studies indicate that the input data for the cod is not very sensitive to the model assumptions made. However, the haddock assessment this year is more uncertain than previously experienced. This is because the 2003 spring survey indices by most year classes are relatively high in relation to information in previous surveys. By keeping to the rigid assumption of constant natural (unaccounted) mortality and fixed catchability in the survey it was difficult to account fully for the residuals in any of the haddock runs, irrespective of software name or model specifications.

The selection of the program to use as the basis of the "final run" for the Icelandic cod and haddock was based on statistical integrity of the methods used as well as on the basis of convenience. The current norm within ICES is to take the assessment through four software packages generating i) historical assessment, ii) recruitment estimates, iii) shortterm predictions, and iv) medium-term prediction. Using statistical catch-at-age one can obtain these estimates within the same "package". As the estimates of different age groups are correlated one can also ignore problems related to selecting estimates of some age groups from one model and other age groups from another model.

### 1.5 Precautionary reference points

No major evaluation of reference points has been made since 1999. The Working Group recognised that some existing reference points may in some cases be inappropriate. Given the management regime in effect in the Faroese demersal fisheries, the reference points for the three main species, cod, haddock, and saithe, should be re-evaluated at the same time. Until more appropriate reference points are identified and adopted, the existing ones could continue to be used, albeit with some flexibility in the formulation of management advice. The catch rule for Icelandic cod has been under revision and it was considered that revision or establishment of reference points for Icelandic haddock and Icelandic saithe should await the result from that analysis. The SG on Precautionary Reference Points for Advice on Fishery Management (SGPRP - February 2003) suggested new $\mathbf{B}_{\text {lim }}$ points for some of the stocks assessed by this Working Group. Considering that ACFM is unlikely to redefine and use new reference points as a basis for advice in the year 2004, the Working Group decided to postpone further work on the issue related to ToRi.

### 1.6 The road ahead

The newly available software at the ICES Secretariat to produce standard graphs for the Working Group and ACFM reports suffers from an antiquated philosophy and needs to be automated so as to minimise user intervention. The user should have to invoke only one program in one location, and she should be directed by the program through the absolute minimum number of steps that cannot be fully automated once file names have been standardised. One important thing to keep in mind in the design phase is that many of the Working Group members may only use this package once a year.

As indicated above, the WG considered assessments using several different approaches for several of the stocks. The WG found it considerably easier to inspect peculiarities of the data using the alternate approaches than using the standard software. The Lowestoft software suite is based on the philosophy of trying to produce the best possible assessment under average conditions, and to minimise the probability of widely erroneous assessments. Yet, there remains a considerable margin for variability in the results by modifying the various settings and input data. The WG sees three main problems with the adopted approach:

1. It is an antiquated piece of software that is awkward to use and with limitations that do not take advantage of the latest developments in computer technology as well as in fisheries modelling;
2. The historical stock assessment software as well as the recruitment generator apply a number of stabilising techniques (time-tapering, shrinking of F and/or P ) whose final weight is often based on very ad hoc and nontransparent decisions.
3. Moving results of estimators of one set of age groups to another program to estimate other age groups ignores the fact that estimates of different age groups are often correlated.

The WG recognises the importance of utilising routines/programs that have been tested for errors/bugs. It also recognises that the standard software package has been and still is, when properly used, a very valuable tool to come up with reasonably sound estimates of stock abundance (most alternative approaches provided results similar to, or consistent with the standard package). However, the WG considers that a more flexible approach to assessment software, in the spirit of the C++ or R software, would be superior to the current insistence on "black box" assessment software. In the proposed spirits, basic individual independent functions/libraries for calculating VPA, Y/R, stock projections would be used by well-trained users.

A simple, although somewhat naive, illustration of the use of the modern methods and computer power to improve the stock assessment process is as follows. The plot below shows a point cloud of 1000 bootstrapped estimates of scaled terminal F and SSB from an unspecified NWWG stock based on a statistical catch-at-age model. Superimposed are the point estimator values (open, larger circles) from different XSA runs that were considered by the WG when selecting "the final run".


Based on the results from the bootstrap estimates it could be argued that the different XSA runs are within the noise in the input data and thus more or less equally likely. If only the XSA point estimators are available, the noise in the data are hidden, and working groups may spend considerable, but meaningless time in trying to justify one run and its settings over the others.

Another example is the continuous use of point estimator in the advice without carrying any information about uncertainty in different stocks. On the left is a comparison of the cumulative bootstrap values of the scaled terminal biomass estimates from two stocks that were assessed by the Working Group. Although the cumulative values should not indicate true significance it is indicative that the noise in the data of one stock is around double that of the other.

It is neither being claimed here that the above examples are something novice nor that they will provide us with a better point estimators. They are intended to show how information that is available in the input data can, by applying simple statistical principles, be of help in improving the assessment and advisory process - something that is hidden and not dealt with in the current ICES approach.

It is the opinion of the WG that the limitations of the standard software packages used by ICES supersedes the issue of software approval. The fast development in the field, particularly in dealing with the uncertainty in the assessments requires that heavy emphasis should be put on certifying the assessor, not just the software. Courses should be designed such that they teach the principles, with emphasis on "how to do it" and not "what to do". Training WG participants in the usage of a particular piece of antiquated software is at best a waste of time for those participating in the training. At worst, it contributes to wasting the time of assessment WG members by being an impediment to more transparent and more modern approaches.

### 2.1 General Trends in Demersal Fisheries in the Faroe Area

The fishery in Faroese waters is a multi-fleet and multi-species fishery. Figure 2.1.1 gives a summary of the 2003 assessments of the stocks of Faroe Plateau cod, Faroe haddock, and Faroe saithe, and Figure 2.1.2 shows the total yield of these stocks.

Fishing mortality on Faroe Plateau cod, Faroe haddock, and Faroe saithe has followed different trends for the three species since the early 1960s (Figure 2.1.1). Fishing mortality for cod and haddock declined steadily from 1961 to the early 1970s, but thereafter evolved differently. For cod, fishing mortality increased and has oscillated around a mean of about $\mathrm{F}=0.50$ since 1974 , with a substantial decrease in the early 1990 s when productivity was lower. For haddock F remained relatively low (between 0.20 and 0.30 ) until the late 1990s when it appears to have increased to pre-1973 values. For saithe, F increased regularly from 1961 to the late 1980s, reaching peak values in the early 1990s, but it appears to have decreased until 1998, with some variability since. When combined in an overall index of exploitation (yield over SSB), the ratio is remarkably stable around 0.30 from 1961 to 1981 (Figure 2.1.2), but since then it has shown larger fluctuations, exceeding 0.55 in 1991. This index of overall exploitation has steadily increased in recent years from slightly less than 0.32 in 1997 to about 0.45 in 2002, during a period in which effort was meant to be constant.

The SSB for cod shows four cycles (Figure 2.1.3) and possibly the beginning of a fifth one, the SSB for saithe two and a half, and the SSB for haddock, three with possibly the beginning of a fourth one. The haddock SSB appears to lag that of cod by 2 years ( $\mathrm{r}=0.82$ ). No such lags are clearly evident for saithe. When added together (Figure 2.1.4), the total SSB increases from 1961 to 1977, then it declines almost steadily until 1992, except for a brief increasing period from 1983 to 1985. SSB has shown a relatively steady increasing trend since then. In 2002, all species are increasing.

Haddock shows the largest recruitment variability (Figure 2.1.5). There is a more than 60 -fold difference between the smallest year class ( 1.8 million) and the largest one ( 110 millions). Cod shows the next largest variability with a 13 -fold difference between the smallest year class ( 3.7 millions) and the largest one ( 48 millions). Saithe shows a 10 -fold difference between the smallest year class ( 8.4 millions) and the largest one ( 80 millions). The recruitment of cod does not show any particular feature other than the string of small year classes during most of the 1980s. Haddock shows sustained recruitment for the 1959 to 1976 year classes, but from 1977 to 1992, only the 1983-1985 year classes were of average size. All the others were much smaller than average. The 1993, 1994, and 1999 year classes are strong. Saithe recruitment increases regularly from the 1958 year class to the 1966 year class and then decreases similarly regularly until the 1975 year class. Recruitment patterns since then have not been so clearly cyclical.

During the 1980s the Faroese authorities have attempted to regulate the fishery and the investment in fishing vessels. In 1987 a system of fishing licenses was introduced. The fishery also has been regulated by technical measures such as legislation on the mesh size, permanent and temporarily area closures, import ban on fishing vessels, and a programme of buying back fishing licenses. Mesh size regulations and closed areas are still enforced.

In March 1994 the Faroese Parliament passed a law on the regulation of fisheries within the EEZ. This law introduced quotas for 5 demersal stocks, including the Faroe Plateau and the Faroe Bank cod, Faroe haddock, Faroe saithe, and redfish. The quotas were allocated to each fleet category by percentage of the total quota and then equally divided between all vessels in each category.

The fishing year starts 1 September and ends 31 August the following year.

### 2.1.1 The management system implemented in 1996

The catch quota management system introduced in the Faroese fisheries in 1994 was met with considerable criticism and it resulted in at least some fleets misreporting substantial portions of their catches. As a result of the dissatisfaction with the catch quota management system, the Faroese Parliament has adopted a law stipulating that the quota system would end as of May 31, 1996. In addition, the Faroese government has developed, in close cooperation with the fishing industry, a new system based on within-fleet category individual transferable effort quotas in days. The new system entered into force on 1 June 1996.

The within-fleet category individual transferable effort quotas apply to 1) the longliners less than 110 GRT, the jiggers and the single trawlers less than $400 \mathrm{HP}, 2$ ) the pair trawlers and 3) the longliners greater than 110 GRT. The single trawlers larger than 400 HP do not have effort limitations, but they are not allowed to fish within the 12 n . miles limit
and the areas closed to them as well to the pairtrawlers have increased in area and time. Their harvest of cod and haddock is limited by maximum bycatch allocation of $4 \%$ and $1.75 \%$. In addition, this fleet ( 13 trawlers) in the present fishing year have been permitted to perform directed cod and haddock fisheries and consequently allocated individual catch quotas of cod and haddock of 100 t each. These quotas have not been accounted for in the allocation of days to other fleets. The single trawlers $<400 \mathrm{HP}$ are given special licenses to fish inside 12 n . miles with a bycatch allocation of $30 \%$ cod and $10 \%$ haddock. Holders of individual transferable effort quotas who fish outside an area where cod and haddock are normally found can fish 3 days for each day allocated within the area of normal cod and haddock distribution. One fishing day by longliners less than 110 GRT is considered equivalent to two fishing days for jiggers in the same gear category. Therefore longliners less than 110 GRT (and single trawlers $<400 \mathrm{HP}$ ) could double their allocation by converting to jigging. Figure 2.1.6 gives an overview of the different area regulations.

The effort quotas are transferable within gear categories. The allocations of number of fishing days by fleet categories was made such that together with other regulations of the fishery they should result in average fishing mortalities on each of the 3 stocks of 0.45 corresponding to average annual catches of $33 \%$ of the exploitable stocks in numbers. Built into the system is also an assumption that the day system is self-regulatory, because the fishery will move between stocks according to the relative availability of each of them and no stock will be overexploited. Pope (2000) examined changes in stock sizes and price and could not find relationships that would support the hypothesis that the economics of the fishery would prevent overfishing of the stocks by shifting the fishing effort to the most abundant species.

The number of days fished by gear category since 1985, and the number of days by category as stated in the law, are presented in Tables 2.1.1 and 2.1.2.

In addition to the number of days allocated in the law, it is also stated in the law what percentage of total catches of cod, haddock, saithe and redfish, each fleet category on average are allowed to fish. These percentages are as follows:

| Fleet category | Cod | Haddock | Saithe | Redfish |
| :--- | :---: | :--- | :---: | :---: |
| Longliners < 110GRT, jiggers, single trawl. < 400HP | $51 \%$ | $58 \%$ | $17.5 \%$ | $1 \%$ |
| Longliners $>$ 110GRT | $23 \%$ | $28 \%$ |  |  |
| Pairtrawlers | $21 \%$ | $10.25 \%$ | $69 \%$ | $8.5 \%$ |
| Single trawlers > 400 HP | $4 \%$ | $1.75 \%$ | $13 \%$ | $90.5 \%$ |
| Others | $1 \%$ | $2 \%$ | $0.5 \%$ | $0.5 \%$ |

Technical measures such as area closures during the spawning periods, to protect juveniles and young fish and mesh size regulations are also in effect (Figure 2.1.6).

Table 2.1.1 Number of fishing days used by various fleet groups in Wb1 $1985-95$ and 1998 -02. For other fleets there are no effort lirnitations Catches of cod, haddock saithe and redfish are regulated by the by-catch percentages given in section 2.1.1. In addition there are special fisheries regulated by licenses and gear festrictions. (This is the real number of days fishing not affected by doubling or tripling of days by changing areas/gears)

| Year | Longliner 0-110 GRT, jiggers, trawlers < 400 HF | Longliners > 110 GRT | Pairtrawlers > 400 HP |
| :---: | :---: | :---: | :---: |
| 1985 | 13449 | 2973 | 8582 |
| 1986 | 11399 | 2176 | 11006 |
| 1987 | 11554 | 2915 | 11860 |
| 1988 | 20736 | 3203 | 12060 |
| 1989 | 28750 | 3369 | 10302 |
| 1990 | 28373 | 3521 | 12935 |
| 1991 | 29420 | 3573 | 13703 |
| 1992 | 23762 | 2892 | 11228 |
| 1993 | 19170 | 2046 | 9186 |
| 1994 | 25291 | 2925 | 8347 |
| 1905 | 33760 | 3659 | 9346 |
| Averager (85-95) | 22333 | 3023 | 10778 |
| 1998 | 23971 | 2519 | 6209 |
| 1999 | 21040 | 2428 | 7135 |
| 2000 | 24820 | 2414 | 7167 |
| 2001 | 29560 | 2512 | 6771 |
| 2002 | 30333 | 2680 | 6749 |
| Averagepe-01) | 29945 | 2511 | 6806 |

Table 2.1.2
Wunber of allocated days for each fleet group since the new management scheme was adopted and number of licenses per fleet

|  | Fleets | 1996/1997 | 1997/1998 | 1998/1999 | 1999/2000 | 2000/2001 | 2001/2002 | 2002/2003 | No. of licenses |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Group 1 | Single trawlers > 400 HP | Regulated by area and by-catch limitations |  |  |  |  |  |  | 13 |
| Group 2 | Pair trawlers > 400 HP | 8225 | 7199 | 6839 | 6839 | 6839 | 6839 | 6771 | 31 |
| Group 3 | Longliners > 110 GRT | 3040 | 2660 | 2527 | 2527 | 2527 | 2527 | 2502 | 19 |
| Group 4 | Longliners and jiggers 15-110 GRT, single trawlers < 400 HP | 9300 | 9328 | 8861 | 8861 | 8861 | 8861 | 8772 | 106 |
| Group 5 | Longliners and jiggers $<15 \mathrm{GRT}$ | 22000 | 23625 | 22444 | 22444 | 22444 | 22444 | 22220 | 696 |



Figure 2.1.1: Faroe demersal stocks fishing mortality.


Figure 2.1.3: Faroe demersal stocks individual SSB trends.


## Figure 2.1.5: Faroe demersal recruitment time series.



Figure 2.1.2: Faroe demersal stocks total landings and overall yield/SSB ratios.


Figure 2.1.4: Faroe demersal stocks spawning stock biomasses.


Areas inside the 12 nm zone closed year round

| Area | Period |
| :---: | :---: |
| a | 1 jan- 31 des |
| aa | 1 jun-31 aug |
| b | 20 jan-1 mar |
| c | 1 jan- 31 des |
| d | 1 jan- 31 des |
| e | $1 \mathrm{apr}-31$ jan |
| f | 1 jan- 31 des |
| g | 1 jan- 31 des |
| h | 1 jan- 31 des |
| i | 1 jan- 31 des |
| j | 1 jan- 31 des |
| k | 1 jan- 31 des |
| 1 | 1 jan- 31 des |
| m | 1 feb- 1 jun |
| n | 31 jan-1 apr |
| o | 1 jan- 31 des |
| p | 1 jan- 31 des |
| r | 1 jan- 31 des |
| S | 1 jan- 31 des |

Figure 2.1.6 Fishing area regulations in Division Vb. Allocation of fishing days applies to the area inside the outer thick line on the Faroe Plateau. Holders of effort quotas who fish outside this line can triple their numbers of days. Longliners larger than 110 GRT are not allowed to fish inside the inner thick line on the Faroe Plateau. If longliners change from longline to jigging, they can double their number of days. The Faroe Bank shallower than 200 m depths ( a , aa) is regulated separate from the Faroe Plateau. It is closed to trawling and the longline fishery is regulated by individual day quotas.

### 2.2.1

 Stock definitionFaroe Plateau cod is distributed on the entire plateau down to approximately the $500-\mathrm{m}$ depth contour. Tagging experiments show that immigration to other areas is very rare (about $0.1 \%$ of recaptured cod; Strubberg, 1916, 1933; Tåning, 1940, 1943; unpublished data). Cod spawn in February-March at two main spawing grounds north and west of the islands at depths of around $90-120 \mathrm{~m}$. The larvae hatch in April and are carried by the Faroe Shelf residual current (Hansen, 1992) that flows clockwise around the Faroe plateau within the 100- to 130-m isobath (Gaard et al. 1998; Larsen et al., 2002). The fry settle in July-August and occupy the near-shore areas, which normally are covered by dense algae vegetation. In autumn the following year (i.e. as 1 -group), the juvenile cod begin to migrate to deeper waters (usually within the $200-\mathrm{m}$ contour), thus entering the feeding areas of adult cod. They seem to be fully recruited to the fishing grounds as 3 -year-olds. Faroe plateau cod mature as 3-4 years old. The spawning migration seems to start in December-January and ends in May. Cod move gradually to deeper waters when they are growing older. The diet in shallow water ( $<200 \mathrm{~m}$ ) is dominated by sandeels and benthic crustaceans, whereas the diet in deeper water mainly consists of Norway pout, blue whiting and a few species of benthic crustaceans.

### 2.2.2 Trends in landings

The annual landings of Faroe cod (ICES Division Vb ) normally varied between 20 and 40 thousand tonnes during the last century. English and Scottish vessels took the majority of the catches up to the 1950s. Thereafter their part of the catches declined gradually, and when the Faroe Islands established the 200 nm EEZ in 1977, the vast majority of the catch was taken by Faroese vessels. From 1965 there have been separate catch figures for Faroe Plateau (ICES Division Vb 1 ) and Faroe Bank (ICES Division Vb2).

The relatively high recruitment in 1980-1983 allowed a good fishery for cod in the period 1983 to 1986 when landings some years reached almost 40000 t . Landings decreased afterwards to only 6000 tonnes in 1993, the lowest on record. In 1995 the officially reported landings increased to slightly above 19000 t . Information from the fishing industry indicated misreporting in the order of 3330 t ( 3000 t gutted weight) for 1995 which were added to the officially reported landings in Table 2.2.1.2. Misreporting is not suspected to have been a problem afterwards. Landings increased spectacularly in 1996, to above 40000 t , the highest value during the 1961 to 2000 time period. This increase is believed to be due to a combination of increased stock size, increased availability, and increased effective fishing effort as a result of the new management system introduced June 1, 1996.

In recent years, statistics for the Faroese fishery in that part of Subdivision IIa (Figure 2.2 ) which is within the Faroese EEZ, have become available. It is expected that these are taken from the Faroe Plateau area so they are included in the total used in the assessment in Table 2.2.1.2 under the row labelled "Used in the assessment". No information on the Faroese landings from IIa were available for 1993-1996, however. The French landings of Faroe Plateau cod in 1989 and 1990 as reported to the Faroese authorities are also included. Scottish catches 1991-1999 reported from the Faroe Bank (Vb2) were in the 2001 assessment moved to the Faroe Plateau (Vb1), by advice from the Faroese Coastal Guard.

Since the introduction of the EEZ, the Faroe Plateau cod has almost entirely been exploited by the Faroese fishing fleets. In recent years, the longliners and the pair trawlers have usually taken most of the catches. Since autumn 1999 single trawlers $>400 \mathrm{HP}$ have increased their share of the total catches considerably as a result of a special quota (in tonnes, not fishing days) allocated to them in shallow water ( $<200 \mathrm{~m}$ ) on a half-year basis (September 1 and March 1).

The nominal landings of cod (1986-2002) from the Faroe Plateau by nations as officially reported to ICES, are given in Table 2.2.1.1. Table 2.2.1.2 shows the figures used in the assessment. In 2002, the catches exceeded 40 thousand tonnes, which is the normal maximum. Table 2.2.1.3 shows the landings for the most important fleet categories.

### 2.2.3 Catch-at-age

The sampling strategy is to have length, length-age, and length-weight samples from all major gears during three periods: January-April, May-August and September-December. In the period 1985-1995, the year was split into four periods: January-March, April-June, July-September, and October-December. When sampling was insufficient, lengthage and length-weight samples were borrowed from similar fleets in the same time period. Length measurements were, if possible, not borrowed.

Landings-at-age were updated to account for a change in the nominal landings for 2001. Landings-at-age for 2002 are provided for the Faroese fishery in Table 2.2.2.1. Faroese landings from most of the fleet categories were sampled (see
text table below). Landings-at-age for the fleets covered by the sampling scheme were calculated from the age composition in each fleet category and raised by their respective landings. The age composition of the combined Faroese landings was used to raise the foreign landings prior to 1998 when the age composition of the corresponding Faroese fleets were used. Landings-at-age from 1961 to 2002 are shown in Table 2.2.2.2. Catch curves are shown in Fig. 2.2.2.1. They show atypical patterns in 1996 and to some extent in 2001, when there appears to be an increase over the previous year for ages where a decrease would normally have been expected. This could be due to catchability for longliners depending on fish growth, causing atypical catch curves for longliners.

Samples from commercial fleets in 2002.

| Fleet | Size | Samples | Length | Otoliths | Weights |
| :--- | :--- | ---: | ---: | ---: | ---: |
| Open boats |  | 33 | 6,485 | 660 | 420 |
| Longliners | $<100$ GRT | 75 | 15,118 | 1,440 | 600 |
| Longliners | $>100$ GRT | 67 | 12,764 | 1,740 | 960 |
| Jiggers | 5 | 907 | 180 | 180 |  |
| Sing. trawlers | $<400 \mathrm{HP}$ | 13 | 2,619 | 296 | 236 |
| Sing. trawlers | $400-1000 ~ H P$ | 27 | 5,371 | 598 | 179 |
| Sing. trawlers | $>1000 \mathrm{HP}$ | 22 | 3,943 | 420 | 240 |
| Pair trawlers | $<1000 \mathrm{HP}$ | 10 | 1,830 | 300 | 240 |
| Pair trawlers | $>1000 \mathrm{HP}$ | 62 | 12,711 | 1,200 | 1,200 |
| Total |  | 281 | 55,263 | 6,174 | 3,835 |

### 2.2.4 Mean weight-at-age

Mean weight-at-age data for 1961-2002 are provided for the Faroese fishery in Table 2.2.3.1. These were calculated using the length/weight relationship based on individual length/weight measurements of samples from the landings. The sum-of-products-check for 2002 showed a discrepancy of less than $1 \%$.

Figure 2.2.3.1 shows the mean weight-at-age for 1961 to 2002. The weights increased from 1998 to high levels in 2000, but have decreased since.

### 2.2.5 Maturity-at-age

The proportion of mature cod by age during the Faroese groundfish surveys carried out during the spawning period (March) are given in Table 2.2.4.1 (1961-2002) and shown in Figure 2.2.4.1 (1983-2003). The average maturity-atage for 1983 to 1996 were used in years prior to 1983. The values for 1994-1997 are revised (Working Document 14) in connection with the correction of the spring groundfish survey, but values prior to 1994 are not changed. The working document deals with the correction of maturities for 1994-1997. On about half of the stations, many fish were incorrectly classified as "maturing", which is highly unlikely for a spring spawner. They were reclassified as "immature" or "spent" according to criteria derived from the years 1999-2002, which data were assumed to be correct. The maturities were calculated in the same way as previously: pooling all fish with information on age and maturity and obtaining the proportion mature directly. Full maturity is generally reached at age 5 or 6 , but considerable changes have been observed in the proportion mature for younger ages between years.

### 2.2.6 Groundfish surveys

The spring groundfish surveys in Faroese waters with the research vessel Magnus Heinason were initiated in 1983. Up to 1991 three cruises per year were conducted between February and the end of March, with 50 stations per cruise selected each year based on random stratified sampling (by depth) and on general knowledge of the distribution of fish in the area. In 1992 the period was shortened by dropping the first cruise and one third of the 1991-stations were used as fixed stations. Since 1993 all stations are fixed stations. The standard abundance estimate is the stratified mean catch per hour in numbers-at-age calculated using smoothed age/length keys. The same strata were used as in the summer survey and calculated in the same way (see below). All cod less than 25 cm were set to 1 year old. The calculation of the age-disaggregated CPUEs from the spring survey should be regarded as preliminary, because the time was limited.

The overall mean catch of cod per unit effort (scaled down to $\mathrm{kg} / 15 \mathrm{~min}$. trawling) 1983-2003 is given in Figure 2.2.5.1 and catch curves in Figure 2.2.5.2. The CPUE increased substantially in 1995 and remained high up to 1998. The CPUE decreased from 2002 to 2003. Normally the stratified mean catch per trawl hour increases for the first $4-5$ years of life
of a year class, and decreases afterwards. From 1994 to 1995, however, there was an increase for all year classes, possibly because of increased availability. A more normal pattern was observed from 1996-2003.

In 1996, a new summer (august-september) groundfish survey was initiated, having 200 fixed stations distributed within the $500-\mathrm{m}$ contour of the Faroe Plateau. Half of the stations were the same as in the spring survey. The overall mean catch of cod per unit effort ( $\mathrm{kg} /$ trawl hour) 1996-2002 is shown in Figure 2.2.5.1, and catch curves in Figure 2.2.5.3. The catch curves show that the fish are fully recruited to the survey gear at an age of 3 or 4 .

The abundance index was calculated as the stratified mean number of cod-at-age. The age-length key was based on otolith samples pooled for all stations since there seemed to be a homogenious size-at-age by strata and depth. Due to incomplete otolith samples for the youngest age groups, all cod less than 15 cm were considered being 0 years and between $15-34 \mathrm{~cm} 1$ year. Since the age-length key was the same for all strata, a mean length distribution was calculated by stratum and the overall length distribution was calculated as the mean length distribution for all strata weighted by stratum area. Having this length distribution and the age-length key, the number of fish at age per station was calculated, and scaled up to 200 stations.

The age-disaggregated CPUEs for the summer series are slightly modified compared to last year. Last year a few hauls were ignored in the calculations due to technical problems.

### 2.2.7 Stock assessment

### 2.2.7.1 Tuning and estimates of fishing mortality

The two tuning series used in NWWG 1998, the single trawlers 400-1000 HP and longliners $>100$ GRT both with fishing effort measured in days were replaced in NWWG 1999 by two tuning series based on logbook data for five longliners > 100 GRT and eight pair trawlers > 1000 HP. In these series, effort was measured in 1000 hooks for the longliners and trawl hours for the pair trawlers. Both tuning series are shown in Figure 2.2.6.1.1 ( $\mathrm{kg} / 1000$ hooks and $\mathrm{kg} /$ hour $)$. The two series show very similar trends for most of the years. Effort standardized catch curves are shown in Figure 2.2.6.1.2 (Cuba trawlers) and Figure 2.2.6.1.3 (longliners). The NWWG 2002 decided to use only the summer groundfish survey as tuning series in the 2002 assessment (see last year's report, ICES, 2002), and this procedure was repeated in this years assessment.

Information about the longliners and Cuba trawlers is found in last year's report (ICES, 2002). The criteria for selecting settings or hauls for CPUE-calculations is changed this year: Instead of using the whole year (January-December), the period February-May was excluded in order to avoid the spawning migration and spawning of cod. During the spawning migration and spawning, mature cod are less accessible (longliners) or not accessible at all (Cuba trawlers). In addition CPUEs may not be an appropriate estimate of stock biomass when the fish are moving and/or are densely aggregated.

Before choosing the final XSA run, four XSA runs were considered: 1) same settings as last year, i.e. the summer survey alone, 2) Cuba trawlers only, 3) longliners only, 4) spring survey only. The diagnostics for the commercial tuning series and the spring survey were poorer than the summer survey. The $\operatorname{logQ}$ residual of the summer survey is shown in Fig. 2.2.6.1.4. The longliner tuning series seemed to have an important deficiency, since the catchability was dependent on the growth rate of cod (Figure 2.2.6.1.5 in last year's report (ICES, 2002)). This suggests that cod preference for longline bait depends on natural food availability. Age 2 was removed from the Cuba series, because they were only expected to catch 2 -year-old cod when the specific year class was large, i.e. they were overestimating large year classes and underestimating poor year classes. The CPUE index of mature or partially mature fish (age 3+) from the spring survey was considered to be poorly defined, because a few very large catches on the spawning grounds had a large influence on the index. Thus, the spring survey was not used in this year's assessment. The results from three XSA runs (summer survey, longliners and Cuba trawlers) are presented in Figure 2.2.6.1.6. The overall picture was the same for all three XSA runs.

For the longliners, only 30 iterations were used. If more iterations were performed, a divergence in fishing mortality was observed in the past (which was unexpected) (see comments on the assessment).

The results from the retrospective analysis of the XSA (Figure 2.2.6.1.5) show that there is a tendency to overestimate fishing mortality and underestimate recruitment, stock biomass and spawning stock biomass.

The estimated fishing mortalities are shown in Tables 2.2.6.1.3 and 2.2.6.1.5 and Figure 2.2.6.1.7. The average F for age groups 3 to 7 in 2002 is estimated at 0.85 , considerably higher than $\mathbf{F}_{\max }=0.48$, but this is due to anomalously large
fishing mortalities on ages 7 (1.8) and 8 (1.6). It therefore gives an erroneous impression of the fishing pressure, but is nevertheless presented for consistency with previous assessments. The ratio yield and exploitable biomass, defined as ages 3 and older biomass ( $\mathrm{Y} / \mathrm{B} 3+$ ), is considered a more reliable indication of the fishing pressure on the stock (Fig. 2.2.6.1.8). The ratio was high during the 1980s and dropped considerably during the poor state of the cod stock in the beginning of the 1990s. It has remained high and stable (between 0.34 and 0.37 ) after the effort management system was introduced in 1996.

### 2.2.7.2 Stock estimates and recruitment

The stock size in numbers is given in Tables 2.2.6.1.4. A summary of the VPA, with recruitment, biomass and fishing mortality estimates is given in Table 2.2.6.1.5 and in Figure 2.2.6.1.7. The stock-recruitment relationship is presented in Figure 2.2.6.2.1.

The assessment shows the poor recruitment for the 1984 to 1991 year classes, and the strong 1992 and 1993 year classes. Due to the continuous poor recruitment from 1984 to 1991 and the high fishing mortalities, the spawning stock biomass declined steadily from 1983 to 1992 when it was the lowest on record at 22000 t . It increased sharply to above 80000 t in 1996 and 1997 before, declining to a level of about 48000 t in 1999. The 1998 year class is above average strength and the 1999 year class is in the current assessment estimated to be as strong as the highest observed (1982 year class: 47 millions). The 2000 year class is estimated to be above average strength.

### 2.2.8 Predictions of catch and biomass

### 2.2.8.1 Short-term prediction

In order to estimate the strength of the year classes 2000-2002, the RCT3 program was used. Table 2.2.6.1.6 shows the input values. The indices used were: 0 -group survey, 1 -group from summer survey (SS1y), 2-group from spring survey (SP2y), 2-group from summer survey (SS2y) and a new index of primary production (P.Prod). Steingrund and Gaard (in submission) have found a positive relationship between an index of primary production and the recruitment of 2-yearold cod the following year. The long time span (1964-2002) was used in order to reduce the influence of the poor recruitment in the period 1987-1991 on the VPA mean. The problem with a few large catches dominating the CPUE index from the spring survey is not supposed to be any problem for immature fish (age 2).

The input data for the short-term prediction are given in Table 2.2.7.1.1. The estimate of year classes 2001-2003 was taken from the RCT3 run. The 2003 year class was estimated as the geometric mean for the period 1961-2002. Estimates of stock size (ages 3+) were taken directly from the VPA stock numbers. The exploitation pattern was estimated as the average fishing mortality for 2000-2002 (not rescaled to 2002 values), omitting high fishing mortalities on ages 7-8 in 2002. The weights-at-age in the catches in 2003 were estimated from the weights in the spring survey. Regression analyses were made between weights in the catches and in the spring survey, and the weights in the catches in 2003 were predicted from those relationships. The weights in the catches in 2004 were set to the values in 2002, since an increase was expected. It was not clear whether the weights in 2005 would increase or decrease, so the values for 2002 were used for 2005 . The proportion mature in 2003 was set to the 2003 values from the spring groundfish survey, and for 2004-2005 to the average values for 2001-2003.

Table 2.2.7.1.2 shows that the landings in 2003 are expected to be 32000 tonnes if the fishing mortality is on the average level of 2000-2002 (excluding high F -values for ages 7-8 in 2002). The spawning stock biomass is expected to be 66000 tonnes in 2003, 82000 tonnes in 2004 and eventually 70000 tonnes in 2005. The VPA suggest that the 1999 year class is as high as the highest observed (1982 year class at 47 millions).

### 2.2.8.2 Biological reference points

In 1998, ACFM set $\mathbf{B}_{\text {lim }}$ equal to the lowest observed SSB , about 21000 t and proposed that $\mathbf{B}_{\mathrm{pa}}$ be set at 40000 t based on $\mathbf{B}_{\mathrm{pa}}=\mathrm{Blime}^{.645} .$. assuming a of about 0.40 to account for the relatively large uncertainties in the assessment. ACFM further proposed that $\mathbf{F}_{\mathrm{pa}}$ be set at 0.35 , more than twice $\mathbf{F}_{0.1}$, about equal to $\mathbf{F}_{\text {max }}$ and $\mathbf{F}_{\text {med }}$ and at the low end of the range of previously estimated $\mathbf{F}_{\text {MSY }}$, from 0.33 (Stefansson and Bell, WD prepared for the SGPAFM) to 0.56 (NWWG, 1997). In previous years, MBAL was considered to be 52000 t . Over the period covered by the assessment, fishing mortality has been equal to or less than this proposed $\mathbf{F}_{\mathrm{pa}}$ in only 6 of 40 years of available data. This suggests that $\mathbf{F}_{\mathrm{pa}}=$ 0.35 may be overly conservative. The updated assessment indicates an $\mathbf{F}_{\text {med }}=0.48, \mathbf{F}_{0.1}=0.27$ and $\mathbf{F}_{\max }=0.48 . \mathbf{F}_{\mathrm{pa}}$ could therefore be set in the order of $\mathbf{F}_{\text {med }}=0.48$. Following the logic used to set $\mathbf{B}_{\mathrm{pa}}, \mathbf{F}_{\text {lim }}$ was set at $\mathbf{F}_{\text {lim }}=$ Fpae1 ${ }^{645}$..that is, $\mathbf{F}_{\text {lim }}=0.68$. The Working Group suggests that the $\mathbf{F}_{\text {lim }}$ be adjusted, using of about 0.40 , according to the review of $\mathbf{F}_{\mathrm{pa}}$, that is $\mathbf{F}_{\text {lim }}=0.90$. The highest fishing mortality in the history of the fishery is in the terminal year (0.85), but this is due to anomalously high values at age 7 and 8 in 2002. The next largest fishing mortality is 0.76 . The Working Group
agrees with the SG on Precautionary Reference Points for Advice on Fishery Management (SGPRP - February 2003) that there is no basis to change the existing $\mathbf{B}_{\mathrm{lim}}$ for Faroe cod.

The stock trajectory with respect to existing reference points is illustrated in Figure 2.2.7.2.1.

### 2.2.8.3 Medium-term prediction

Medium-term 20 years prediction were done in the 2001 assessment (ICES 2001). It was not repeated this year.

### 2.2.8.4 Long-term prediction

The input data for the yield-per-recruit calculations (long-term predictions) are given in Table 2.2.7.4.1. The exploitation pattern (not rescaled to 2002 values) and weight-at-age were set to the average values for 1961-2002. The proportion mature was set to the average for 1983-2003.

The output from the yield-per-recruit calculations is shown in Table 2.2.7.4.2. and in Figure 2.2.7.4.1. $\mathbf{F}_{0.1}$ was calculated as 0.27 and $\mathbf{F}_{\text {max }}$ as 0.48 . The present average fishing mortality (F3-7) in 2002 of 0.85 is substantially above $\mathbf{F}_{\text {max }}$ and $\mathbf{F}_{\text {med }}=0.48$ (Figure 2.2.7.2.1). (The program that calculates the long-term prediction has, incorrectly, interchanged the values of $\mathbf{F}_{\text {low }}$ and $\mathbf{F}_{\text {high }}$ ).

### 2.2.9 Management considerations

The management system with individual transferable days introduced in 1996 had as an objective to decrease fishing mortality. The current assessment shows that instead, fishing mortality increased from 0.32 in 1995 to 0.69 in 1996. The WG report for 2000 describes the scope for changes in catchability and how they could account for such increases in fishing mortality, and it also reports on an external review of the scientific basis for the initial allocation of fishing days and of the method to calculate probability profiles for expected fishing mortalities, given the possible utilisation of the allocated fishing days (Pope 2000).

Given the recent history, however, fishing mortality in future years is expected to be above the proposed $\mathbf{F}_{\mathrm{pa}}$ of $\mathrm{F}=0.35$ unless the number of days are reduced substantially.

For reference purposes the number of days allocated to each fleet category are given in the table below:

| Gear <br> LL $<110$ | Allocation <br> 8861 | Optional change <br> There are 8861 days to be shared/chosen to be fished either by longlining <br> $(<100)$, jigging or trawling $(<400 \mathrm{hp})$ |
| :--- | :--- | :--- |
| ST $<400$ | 0 | There are 8861 days to be shared/chosen to be fished either by longlining <br> $(<100)$, jigging or trawling $(<400 \mathrm{hp})$ |
| ST400-1000 | 0 | No effort limitation, assumed to catch less than $4 \%$ cod. |
| ST $>1000$ | 0 | No effort limitation, assumed to catch less than $4 \%$ cod. |
| PT $>400$ | 6839 | 2527 |
| LL $>110$ | 22444 |  |
| OPEN |  | There are 8861 days to be shared/chosen to be fished either by longlining <br> $(<100)$, jigging or trawling $(<400 \mathrm{hp})$ |

In addition to the effort control, the fleets are supposed to be constrained to a pre-agreed species composition in the catch as indicated in the table below:

| Groups of fleets | Fleet | Cod | Haddock | Saithe | Redfish |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | \% | \% | \% | \% |
| Group 1 | Single trawlers | 4.0 | 1.75 | 13.0 | 90.5 |
| Group 2 | Pair trawlers | 21.0 | 10.25 | 69.0 | 8.5 |
| Group 3 | Longliners > 100 GRT | 23.0 | 28.0 |  |  |
| Group 4 | Longliners and jiggers > 15 GRT | 31.0 | 34.5 | 11.5 | 0.5 |
| Group 5 | Longliners and jiggers < 15 GRT | 20.0 | 23.5 | 6.0 |  |


| Group 6 | Others | 1.0 | 2.0 | 0.5 | 0.5 |
| :---: | :---: | ---: | ---: | ---: | ---: |
|  |  | 100 | 100 | 100 | 100 |

These restrictions do not take into account that several of these fleets are in fact involved in a multispecies fishery and that the actual species composition in the water is unlikely to be exactly the same as in catches under the regulation. The percentages are guidelines only and it is not expected they will result in discarding and misreporting. They are therefore unlikely to jeopardise one of the eventual potential benefits of an effort management system, an improvement in the quality of the information collected from the fisheries.

Management systems based on effort controls are expected to lead to overcapitalisation in the fishing fleets because vessel owners will want to maximise the catch they can harvest with the fishing effort allocation they have received. In the medium to long term, this process will lead to increased fishing efficiency of the fleets and it will be necessary to decrease the total number of fishing days available to be allocated in order not to exert excessive fishing mortality. In extreme cases, effort controls can lead to the fishery being open only for a few days per year as was the case for the Pacific halibut fishery a few years ago, and remains the case for some Pacific herring fisheries off the Coast of British Columbia.

In order to constrain fishing mortality within reasonable limits, it will therefore be necessary to adjust the number of days periodically. For this purpose, there is a need for a mechanism to monitor changes in efficiency, and detailed information on the activities of the fleets, on the physical characteristics of the boats and their equipment should therefore be collected. In the case of Faroe Plateau cod, the results of medium-term simulations presented in ICES 2001 suggest that fishing mortality should be decreased by $25 \%$. This is much more than the $1 \%$ reduction in fishing days for the fishing year 2002-2003.

If the intent of fishery management is to control fishing mortality within some limits, it is important that control be exerted on all fleet components generating fishing mortality. In the Faroes, the main tool to control fishing mortality on cod, haddock and saithe is the effort management system and area closures. Single trawlers larger than 400 HP are limited by fishing area and any possible reduction of fishing days would not affect them. As long as they catch less than $4 \% \operatorname{cod}$ (as is assumed), the mortality they generate may not be problematic. However, since 1999 they have been allocated cod and haddock quotas, and in theory the additional fishing pressure they generate should be compensated by a reduction in the total number of days allocated to other fleets. Although discarding is not believed to be a serious problem in the Faroes, management by catch quotas provides for incentives for such behavior.

From an ecological point of view, it should be an advantage to reduce the fishing mortality. The abundance of large cod would increase and the predation on blue whiting should be higher. Since the majority of the blue whiting stock is growing up outside the Faroe Plateau, and mainly is present in the Faroese area in connection with the spawning migration, an increased predation on blue whiting should represent an influx of energy to the Faroe Plateau ecosystem and eventually a higher catch of cod.

### 2.2.10 Comment on the assessment

New or changed things compared to last year's report: The assessment was done in exactly the same way as last year. The estimation of recent year classes by the RCT3 program now was based on two indices more: The 2 year olds from the spring survey and an index of primary production.

In the preliminary XSA runs, there seemed to be a problem with the XSA run when using the longliners as the only tuning series. When having more than 30 iterations, unexpected deviations in fishing mortalities for some years back in time were obtained (Fig. 2.2.9.1). When stopping at 30 iterations, more likely values were obtained (Fig. 2.2.6.1.6). These problems were not observed when using ADAPT, because the assumptions about the fishing mortalities on the oldest ages is different.

In last year's report, the main problem in the short-term prediction was the uncertainty about the large 1999 year class as well as the year classes 2000-2001, because few indices of year-class strength were available. This years assessment shows that the 1999 year class is as strong as expected, and more indices are available to assess the strength of recent year classes (2000-2002 year classes). In this sense, the short-term prediction this year should be more accurate compared to last year. On the other hand it was difficult to decide on the selection pattern because it varied so much between years. It was decided to use the default year range (last 3 years) and excluding high fishing mortalities on ages 7 and 8 in 2002. Since no trend in the fishing mortalities was detected, no rescaling to the final year was done. The short-term prediction predicted the catch in 2003 to be about 32000 tonnes. This is less than expected, since the landings the first three months in 2003 are similar to the same period in 2002 (and the catch in 2002 was 40000 tonnes). A statistical catch-at-age model (see Icelandic cod) predicted a catch in the range of 37-40 000 tonnes.

The most important change compared to last year's assessment is the perception of fishing mortality and stock size. The F3-7 in 2001 has changed from 0.71 to 0.45 which shows that the advice should not only be based on F3-7. The stock biomass and spawning stock biomass are estimated to be around 10000 t higher in recent years (1999-2001). The changes in the spawning stock biomass in 1994-1995 (reduction of about 10000 t ) comes from the revised maturities. The perception of recruitment is very similar to the 2002 assessment.

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Table 2.2.1.1 Faroe Plateau ( Subdivision Vb1) COD. Nominal catches (tonnes) by countries, 1986-2002, as officially reported to ICES.

|  | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |  | 1993 |  | 1994 | 1995 |  | 1996 | 1997 | 1998 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 8 | 30 | 10 | - | - | - | - |  | - |  | - | - |  | - | - | - |
| Faroe Islands | 34,492 | 21,303 | 22,272 | 20,535 | 12,232 | 8,203 | 5,938 |  | 5,744 |  | 8,724 | 19,079 |  | 39,406 | 33,556 | 23,308 |
| France | 4 | 17 | 17 | - | - | - ${ }^{1}$ | 3 | 2 | 1 | 2 | - | 2 | 2 | $1^{2}$ | - | - |
| Germany | 8 | 12 | 5 | 7 | 24 | 16 | 12 |  | + |  | $2^{2}$ | 2 |  | + | + | - |
| Norway | 83 | 21 | 163 | 285 | 124 | 89 | 39 |  | 57 |  | 36 | 38 |  | 507 | 410 | 405 |
| Greenland | - | - | - | - | - | - | - |  | - |  | - | - |  | - | - | - |
| UK (Engl. and Wales) | - | 8 | - | - | - | 1 | 74 |  | 186 |  | 56 | 43 |  | 126 | $61{ }^{2}$ | $27^{2}$ |
| UK (Scotland) | - | - | - | - | - | - | - |  | - |  | - | - |  | - | - | - |
| United Kingdom | - | - | - | - | - | - | - |  | - |  | - | - |  | - | - | - |
| Total | 34,595 | 21,391 | 22,467 | 20,827 | 12,380 | 8,309 | 6,066 |  | 5,988 |  | 8,818 | 19,164 |  | 40,040 | 34,027 | 23,740 |


|  | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: |
| Denmark | - |  |  |  |
| Faroe Islands | 19,156 |  |  |  |
| France ${ }^{1)}$ | - | 1 | 7* | 20 |
| Germany | 39 | 2 | 9 | $6{ }^{2}$ |
| Norway | 450 | 374 * | 544 * | 732 |
| Greenland |  |  |  |  |
| UK (Engl. and Wales) | $51^{2}$ | $18{ }^{2}$ | $50{ }^{2}$ |  |
| UK (Scotland) | - |  |  |  |
| United Kingdom |  |  |  | 1 |
| Total | 19,696 | 395 | 610 | 758 |

[^0]Table 2.2.1.2 Nominal catch (tonnes) of COD in Subdivision Vb1 (Faroe Plateau) 1986-2002, as used in the assessment.

|  | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Officially reported | 34,595 | 21,391 | 22,467 | 20,827 | 12,380 | 8,309 | 6,066 | 5,988 | 8,818 | 19,164 | 40,040 | 34,027 | 23,740 |
| Faroese catches in IIA within |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Faroe area jurisdiction |  |  | 715 | 1,229 | 1,090 | 351 | 154 |  |  |  |  |  |  |
| Expected misreporting/discard |  |  |  |  |  |  |  |  |  | 3330 |  |  |  |
| French catches as reported |  |  |  |  |  |  |  |  |  |  |  |  |  |
| to Faroese authorities |  |  |  | 12 | 17 |  |  |  |  |  |  |  |  |
| Catches reported as Vb2: |  |  |  |  |  |  |  |  |  |  |  |  |  |
| UK (E/W/NI) |  |  |  |  | - | - | + | 1 | 1 | - | - | - | - |
| UK (Scotland) |  |  |  |  | 205 | 90 | 176 | 118 | 227 | 551 | 382 | 277 | 265 |
| Used in the assessment | 34,595 | 21,391 | 23,182 | 22,068 | 13,487 | 8,750 | 6,396 | 6,107 | 9,046 | 23,045 | 40,422 | 34,304 | 24,005 |
|  | 1999 | 2000 | 2001 | 2002 |  |  |  |  |  |  |  |  |  |
| Officially reported | 19,696 | 395 | 610 | 758 |  |  |  |  |  |  |  |  |  |

Faroese catches in Vb1 $\quad 21,793^{*} \quad 28,511 \quad 39,102 *$

Greenland
26

| Catches reported as Vb2: |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
| UK (E/W/NI) | - | - | - |  |
| UK (Scotland) | 210 | 245 | 288 |  |
| United Kingdom |  |  |  | 273 |
| Used in the assessment | 19,906 | 22,433 | 29,409 | 40,159 |

*) Preliminary

Table 2.2.1.3 Faroe Plateau (Subdivision Vb1) COD. The landings of Faroese fleets (in percents) of total catch.

| Year | Open <br> boats |  | Longliners <100 GRT | Singletrawl $<400 \mathrm{HP}$ | Gill net |  | Jiggers |  | Singletrawl 400-1000 HP | Singletrawl $>1000 \mathrm{HP}$ | Pairtrawl < 1000 HP | $\begin{aligned} & \hline \text { Pairtrawl } \\ & >1000 \mathrm{HP} \end{aligned}$ | Longliners $>100$ GRT | Industrial trawlers | Others | Total <br> Round.weig |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 |  | 9.5 | 15.1 | 5.1 |  | 1.3 |  | 2.9 | 6.2 | 8.5 | 29.6 | 14.9 | 5.1 | 0.4 | 0.41 .3 | 34,492 |
| 1987 |  | 9.9 | 14.8 | 6.2 |  | 0.5 |  | 2.9 | 6.7 | 8.0 | 26.0 | 14.5 | 9.9 | 0.5 | 50.1 | 21,303 |
| 1988 |  | 2.6 | 13.8 | 4.9 |  | 2.6 |  | 7.5 | 7.4 | 6.8 | 25.3 | 15.6 | 12.7 | 0.6 | 60.2 | 22,272 |
| 1989 |  | 4.4 | 29.0 | 5.7 |  | 3.2 |  | 9.3 | 5.7 | 5.5 | 10.5 | 8.3 | 17.7 | 0.7 | $7 \quad 0.0$ | 20,535 |
| 1990 |  | 3.9 | 35.5 | 4.8 |  | 1.4 |  | 8.2 | 3.7 | 4.3 | 7.1 | 10.5 | 19.6 | 0.6 | $6 \quad 0.2$ | 12,232 |
| 1991 |  | 4.3 | 31.6 | 7.1 |  | 2.0 |  | 8.0 | 3.4 | 4.7 | 8.3 | 12.9 | 17.2 | 0.6 | $6 \quad 0.1$ | 8,203 |
| 1992 |  | 2.6 | 26.0 | 6.9 |  | 0.0 |  | 7.0 | 2.2 | 3.6 | 12.0 | 20.8 | 13.4 | 5.0 | 00.4 | 5,938 |
| 1993 |  | 2.2 | 16.0 | 15.4 |  | 0.0 |  | 9.0 | 4.1 | 3.6 | 14.2 | 21.7 | 12.6 | 0.8 | $8 \quad 0.4$ | 5,744 |
| 1994 |  | 3.1 | 13.4 | 9.6 |  | 0.5 |  | 19.2 | 2.7 | 5.3 | 8.3 | 23.7 | 13.7 | 0.5 | 50.1 | 8,724 |
| 1995 |  | 4.2 | 17.9 | 6.5 |  | 0.3 |  | 24.9 | 4.1 | 4.7 | 6.4 | 12.3 | 18.5 | 0.1 | 10.0 | 19,079 |
| 1996 |  | 4.0 | 19.0 | 4.0 |  | 0.0 |  | 20.0 | 3.0 | 2.0 | 8.0 | 19.0 | 21.0 | 0.0 | 0.0 | 39,406 |
| 1997 |  | 3.1 | 28.4 | 4.4 |  | 0.5 |  | 9.8 | 5.1 | 2.9 | 4.8 | 11.3 | 29.7 | 0.0 | 0.1 | 33,556 |
| 1998 |  | 2.4 | 31.2 | 6.0 |  | 1.3 |  | 6.5 | 6.3 | 5.5 | 3.1 | 8.6 | 29.1 | 0.1 | 10.0 | 23,308 |
| 1999 |  | 2.7 | 24.0 | 5.4 |  | 2.3 |  | 5.4 | 5.2 | 11.8 | 6.4 | 14.5 | 21.9 | 0.4 | 40.1 | 19,156 |
| 2000 |  | 2.3 | 19.3 | 9.1 |  | 0.9 |  | 10.5 | 9.6 | 12.7 | 5.7 | 13.9 | 15.7 | 0.1 | 10.1 | 21,793 |
| 2001 |  | 3.7 | 28.3 | 7.4 |  | 0.2 |  | 15.6 | 6.4 | 6.4 | 5.2 | 9.2 | 17.8 | 0.0 | 00 | 28,099 |
| 2002 |  | 3.8 | 32.9 | 5.8 |  | 0.3 |  | 9.9 | 6.7 | 6.6 | 2.5 | 7.2 | 24.4 | 0.0 | 0.0 | 39,102 |

Table 2.2.2.1 Faroe Plateau COD. Catch in numbers-at-age per fleet in 2002. Numbers are in thousands and the catch is in tonnes, round weight.

| AgelFleet | Open boats longline | Open boats jiggers | Longliners < 100 GRT | Jiggers | Single trwl 0-399HP | Single trwl $400-1000 \mathrm{H}$ | Single trwl $\text { H> } 1000 \text { HP }$ | Pair trwl $700-999 \mathrm{HI}$ | $\begin{aligned} & \text { Pair trwl } \\ & \text { II> }>1000 \mathrm{HP} \end{aligned}$ | Longliners $>100 \text { GRT }$ | Gillnetters | Others | Catch-at-age |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 74 | 67 | 1114 | 207 | 26 | 52 | 17 | 7 | 29 | 444 | 0 | 134 | 2171 |
| 3 | 159 | 127 | 2743 | 725 | 353 | 452 | 340 | 140 | 402 | 1700 | 1 | 463 | 7605 |
| 4 | 61 | 64 | 1038 | 351 | 232 | 281 | 210 | 98 | 302 | 665 | 3 | 216 | 3521 |
| 5 | 43 | 31 | 546 | 177 | 133 | 140 | 95 | 32 | 80 | 359 | 2 | 107 | 1745 |
| 6 | 6 | 12 | 160 | 47 | 28 | 30 | 20 | 6 | 16 | 134 | 2 | 30 | 491 |
| 7 | 10 | 11 | 223 | 39 | 20 | 24 | 30 | 5 | 12 | 148 | 2 | 32 | 556 |
| 8 | 7 | 5 | 204 | 32 | 20 | 18 | 6 | 2 | 8 | 100 | 3 | 26 | 431 |
| 9 | 2 | 3 | 89 | 25 | 12 | 15 | 8 | 3 | 12 | 113 | 2 | 19 | 303 |
| 10+ | 0 | 0 | 5 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 7 |
| Sum | 362 | 320 | 6122 | 1603 | 824 | 1013 | 726 | 293 | 861 | 3663 | 15 | 1028 | 16830 |
| G.weight | 637 | 638 | 11175 | 3357 | 1970 | 2262 | 2246 | 838 | 2454 | 8297 | 92 | 2213 | 36179 |



Table 2.2.3.1. Faroe Plateau COD. Catch weight-at-age 1961-2002.



Table 2.2.6.1.1 Faroe Plateau (Subdivision Vb1) COD. Summer survey tuning series (number of individuals per 200 stations).

| FAROE PLATEAU | COD (ICES | SUBDIVISION VB1) |  |  | SU.TXT |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 101 |  |  |  |  |  |  |
| SUMMER SURVEY |  |  |  |  |  |  |
| 19962002 |  |  |  |  |  |  |
| 110.60 .7 |  |  |  |  |  |  |
| 28 |  |  |  |  |  |  |
| $200 \quad 829.7$ | 6317.1 | 3840.5 | 1416.5 | 703 | 244.4 | 51.4 |
| 200566.2 | 1839.8 | 6263.6 | 1597.7 | 179 | 140.3 | 30.2 |
| 200518.4 | 548.4 | 1104.9 | 3517.5 | 973.8 | 53.6 | 37.2 |
| 200372.3 | 1267.1 | 778.8 | 754 | 1298.3 | 256.5 | 38.7 |
| 2001344.3 | 1132.3 | 697.2 | 315.5 | 434.6 | 614.9 | 35.5 |
| 2003375.1 | 2471.4 | 1524.7 | 429.5 | 246.6 | 297.3 | 248.6 |
| 2002289.2 | 5198.4 | 1794.4 | 806.5 | 145.3 | 85.8 | 70.5 |

Table 2.2.6.1.2 Faroe Plateau (Subdivision Vb1) COD. Final XSA run.

## Lowestoft VPA Version 3.1

2/05/2003 17:08
Extended Survivors Analysis

```
COD FAROE PLATEAU (ICES SUBDIVISION Vb1) COD_IND_su
CPUE data from file SU.TXT
Catch data for 42 years. 1961 to 2002. Ages 2 to 9.
    Fleet, First, Last, First, Last, Alpha, Beta
    year, year, age , age
    , 1996, 2002, 2, 8, .600, . 700
Time-series weights :
    Tapered time weighting not applied
Catchability analysis :
    Catchability dependent on stock size for ages < 3
        Regression type = C
        Minimum of 5 points used for regression
        Survivor estimates shrunk to the population mean for ages < 3
    Catchability independent of age for ages >= 6
Terminal population estimation :
Survivor estimates shrunk towards the mean F
of the final 5 years or the 5 oldest ages.
S.E. of the mean to which the estimates are shrunk = 2.000
Minimum standard error for population
estimates derived from each fleet =
. 300
Prior weighting not applied
Tuning had not converged after 30 iterations
Total absolute residual between iterations
29 and 30=.00046
Final year F values
```



```
Iteration 29, .1081, . 2714, . 5399, .9083, . 7050, 1.8313, 1.6077, . 8995
Iteration 30, . 1081, . 2714, . 5399, .9083, . 7050, 1.8314, 1.6078, .8997
Regression weights
        ,1.000,1.000, 1.000, 1.000, 1.000, 1.000, 1.000
Fishing mortalities
    Age, 1996, 1997, 1998, 1999, 2000, 2001, 2002
            2,.030, .035, .081, .102, .121, .099, . . 08
            3, .187, .146, .174, . 279, . 336, . 323, . 271
            4,.447, . 395, . 266, . 317, . 361, .479, . 540
            5, . 783, . 816, . 603, . 339, . 272, . 280, . . 08
            6, . 888, .962, .999, . 646, . 349, . 387, . 705
            7, 1.134, 1.314, .650, 1.061, .504, .760, 1.831
            8, . 868, 1.326, .951, . 610, .770, . 548, 1.608
            9, .833, .882, . 813, . 374, . 135, . 664, . . 000
```


## Table 2.2.6.1.2 (Cont'd)

1
XSA population numbers (Thousands)
YEAR , 2, AGE

9 ,

| 1996 |  | 1 | 3 | 1.41E+04, | 3. | 2.87E+03, | 9. | 2. | $6.79 \mathrm{E}+02$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 |  | $6.50 \mathrm{E}+03$, | 1.04E+04, | 2.27E+04, | 7.39E+03, | 1.18E+03, | $9.66 \mathrm{E}+02$ | $2.56 \mathrm{E}+02$, | 01 |
| 1998 |  | $6.49 \mathrm{E}+03$, | 5.14E+03, | 7.37E+03, | 1.25E+04, | 2.67E+03, | $3.68 \mathrm{E}+02$, | 2.13E+02, | $5.56 \mathrm{E}+01$ |
| 1999 |  | $1.47 \mathrm{E}+04$, | 4.91E+03, | 3.54E+03, | 4.62E+03, | 5.62E+03, | 8.06E+02, | 1.57E+02, | $6.72 \mathrm{E}+01$ |
| 2000 |  | 2.17E+04, | 1.09E+04, | 3.04E+03, | 2.11E+03, | 2.70E+03, | $2.41 \mathrm{E}+03$ | 2.28E+02, | $6.99 \mathrm{E}+0$ |
| 2001 |  | 4.77E+04, | $1.57 \mathrm{E}+04$, | $6.37 \mathrm{E}+03$, | 1.73E+03, | 1.32E+03, | $1.56 \mathrm{E}+03$ | $1.19 \mathrm{E}+03$, | $8.65 \mathrm{E}+0$ |
| 2002 |  | 2.34E+04, | 3.54E+04, | 9.33E+03, | 3.23E+03, | 1.07E+03, | 7.32E+02, | $5.96 \mathrm{E}+02$, | $5.64 \mathrm{E}+02$ |

Estimated population abundance at 1st Jan 2003

```
0.00E+00, 1.72E+04, 2.21E+04, 4.45E+03, 1.07E+03, 4.34E+02, 9.60E+01, 9.77E+01,
```

Taper weighted geometric mean of the VPA populations:
$1.55 \mathrm{E}+04,1.14 \mathrm{E}+04,6.87 \mathrm{E}+03,3.69 \mathrm{E}+03,1.83 \mathrm{E}+03,8.57 \mathrm{E}+02,3.57 \mathrm{E}+02,1.50 \mathrm{E}+02$,
Standard error of the weighted Log(VPA populations) :
. .5711, .5635, .5364, .5367, .5602, .5758, .6331, .7651,

Log-catchability residuals.

Fleet : SUMMER SURVEY

| Age, | 1996, | 1997, | 1998, | 1999, | 2000, | 2001, | 2002 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2, | -.06, | .25, | .19, | -.95, | -.01, | .13, | .45 |
| 3, | .16, | .06, | -.42, | .53, | -.34, | .06, | -.04 |
| 4, | .24, | .21, | -.48, | -.06, | .01, | .13, | -.05 |
| 5 | .78, | .07, | .19, | -.52, | -.65, | -.14, | .27 |
| 6, | .32, | -.11, | .79, | .11, | -.45, | -.27, | -.39 |
| 7, | .51, | .07, | -.35, | .69, | .11, | -.01, | .20 |
| 8, | .02, | -.12, | .03, | .15, | -.21, | -.06, | .06 |

Mean log-catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age , | 3, | 4, | 5, | 6, | 7, |
| ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -6.8697, | -6.4154, | -6.2392, | -6.3190, | -6.3190, |
| S.E (Log q), | .3191, | .2423, | .4913, | .4427, | .3920, |

Regression statistics :
Ages with $q$ dependent on year class strength
Age, Slope, t-value, Intercept, RSquare, No Pts, Reg s.e, Mean Log q
2, 1.02,
-.084,
7.81, .72,
7, .49, -7.85,

Ages with $q$ independent of year class strength and constant w.r.t. time.

Age, Slope , t-value , Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 3, | .99, | .063, | 6.90, | .86, | 7, | .35, | -6.87, |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 4, | .89, | .930, | 6.71, | .93, | 7, | .22, | -6.42, |
| 5, | .86, | .530, | 6.53, | .74, | 7, | .45, | -6.24, |
| 6, | .75, | 1.151, | 6.66, | .81, | 7, | .32, | -6.32, |
| 7, | .89, | .498, | 6.23, | .80, | 7, | .33, | -6.14, |
| 8, | 1.02, | -.257, | 6.35, | .97, | 7, | .13, | -6.34, |

Table 2.2.6.1.2 (Cont'd)


Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year, | s.e, | s.e, | Ratio, |  |  |
| $17216 .$, | .39, | .44, | 3, | 1.110, | .108 |

Age 3 Catchability constant w.r.t. time and dependent on age
Year class $=1999$

SUMMER SURVEY , 22020., .298, .070, .23, 2, .971, .272
F shrinkage mean , 23922., 2.00, , , . 029 , 253

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year, | S.e, | s.e, | Ratio, |  |  |
| $22073 .$, | .29, | .05, | 3, | .169, | .271 |

1
Age 4 Catchability constant w.r.t. time and dependent on age
Year class $=1998$

| Fleet, |  | Estimated, Survivors, | $\begin{aligned} & \text { Int, } \\ & \text { s.e, } \end{aligned}$ | $\begin{aligned} & \text { Ext, } \\ & \text { s.e, } \end{aligned}$ | Var, <br> Ratio, | N, | Scaled, Weights, | $\begin{aligned} & \text { Estimated } \\ & \mathrm{F} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SUMMER SURVEY | , | 4403., | . 211, | . 035, | .17, | 3, | . 979 , | . 544 |
| F shrinkage mean |  | 7211., | 2.00, |  |  |  | . 021, | .366 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year, | S.e, | S.e, | Ratio, |  |  |
| $4450 .$, | .21, | .05, | 4, | .240, | .540 |

Age 5 Catchability constant w.r.t. time and dependent on age
Year class $=1997$

| Fleet, |  | Estimated, Survivors, | Int, s.e, | Ext, s.e, | Var, Ratio, | N, | Scaled, Weights, | $\begin{gathered} \text { Estimated } \\ \mathrm{F} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SUMMER SURVEY |  | 1028., | . 204 , | . 201, | .99, | 4, | . 962 , | 930 |
| F shrinkage mean |  | 2662., | 2.00, |  |  |  | . 038, | . 466 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | Ratio, |  |  |
| $1067 .$, | .21, | .19, | 5, | .923, | .908 |

Table 2.2.6.1.2 (Cont'd)

| Age 6 Catchabili |  | nstant w.r | time | epende | on ag |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year class $=1996$ |  |  |  |  |  |  |  |  |
| Fleet, |  | Estimated, | Int, | Ext, | Var, |  | Scaled, | Estimated |
| , |  | Survivors, | s.e, | s.e, | Ratio, |  | Weights, | F |
| SUMMER SURVEY | , | 433., | .191, | .157, | . 82, | 5, | . 972 , | .706 |
| F shrinkage mean | , | 461., | 2.00, |  |  |  | . 028, | . 675 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | S.e, | Ratio, |  |  |
| $434 .$, | .19, | .14, | 6, | .713, | .705 |


| Age 7 Catchabil |  | nstant w | tim | ge (fis | d at |  | ue for | ge) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year class $=1995$ |  |  |  |  |  |  |  |  |
| Fleet, |  | Estimated, Survivors, | Int, | Ext, s.e, | Var, <br> Ratio |  | Scaled, Weights, | Estimated <br> F |
| SUMMER SURVEY | , | 86. | .183, | . 130, | . 71, |  |  | 1.927 |
| F shrinkage mean |  | 365., | 2.00, |  |  |  | .078, | . 867 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year, | S.e, | S.e, | Ratio, |  |  |
| $96 .$, | .23, | .20, | 7, | .875, | 1.831 |

1
Age 8 Catchability constant w.r.t. time and age (fixed at the value for age) 6
Year class $=1994$


Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | Ratio, |  |  |
| 98., | .21, | .13, | 8, | .631, | 1.608 |

Age 9 Catchability constant w.r.t. time and age (fixed at the value for age) 6
Year class $=1993$


Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| $188 .$, | .22, | .05, | 7, | .251, | .900 |

Table 2.2.6.1.3. Faroe Plateau (Subdivision Vb1) COD. Fishing mortality-at-age.

Run title : COD FAROE PLATEAU (ICES SUBDIVISION Vb1) At 2/05/2003 17:09

Terminal Fs derived using XSA (With F shrinkage)

|  |  | Table YEAR, | $8$ | $\begin{aligned} & \text { Fishing } \\ & \text { 1961, } \end{aligned}$ | $\begin{gathered} \text { mortality } \\ 1962, \end{gathered}$ | (F) at | age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | 2, |  | . 3346 , | . 2701 , |  |  |  |  |  |  |  |  |
|  |  | 3, |  | . 5141, | . 4982 , |  |  |  |  |  |  |  |  |
|  |  | 4, |  | . 4986, | . 4838, |  |  |  |  |  |  |  |  |
|  |  | 5, |  | . 5737, | . 7076 , |  |  |  |  |  |  |  |  |
|  |  | 6 , |  | . 4863, | . 5569, |  |  |  |  |  |  |  |  |
|  |  | 7, |  | . 9566 , | . 3662 , |  |  |  |  |  |  |  |  |
|  |  | 8, |  | . 8116, | . 6826 , |  |  |  |  |  |  |  |  |
|  |  | 9, |  | .6715, | . 5641, |  |  |  |  |  |  |  |  |
| 0 | FBAR | 3-7, |  | .6059, | . 5226, |  |  |  |  |  |  |  |  |
|  |  | Table | 8 | Fishing | mortality | (F) at |  |  |  |  |  |  |  |
|  |  | YEAR, |  | 1963, | 1964, | 1965, | 1966, | 1967, | 1968, | 1969, | 1970, | 1971, | 1972, |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 2, |  |  | . 2534 , | .1086, | .1209, | .0829, | . 0789 , | .1010, | .1099, | . 0530, | .0309, | . 0464 , |
|  | 3, |  |  | . 4138, | . 2997, | . 2518, | . 1969, | . 2389 , | . 2318, | . 3063 , | . 2081, | .1337, | . 1476, |
|  | 4, |  |  | . 5172, | . 4523, | . 4498 , | . 2552 , | . 2687 , | . 3949 , | . 3806 , | . 3654 , | . 2225, | . 2070, |
|  | 5, |  |  | . 5124, | . 5229, | . 5622 , | .4499, | . 3442 , | .5339, | . 4180, | . 3409 , | . 3845 , | . 2497, |
|  | 6 , |  |  | .5405, | . 5659, | . 6604 , | . 5016, | . 5779 , | . 4472 , | .5709, | . 3709 , | . 5572, | . 6058, |
|  | 7, |  |  | . 4879, | . 6677, | . 5305 , | . 9680 , | .5203, | . 7132 , | . 5118, | .6559, | . 4651, | . 4686 , |
| 8, |  |  |  | . 3269 , | . 3531, | .4345, | . 8520 , | 1.0438, | . 3331, | .8457, | . 4208, | . 7528 , | . 2464 , |
| 9, |  |  |  | . 4806 , | . 5164 , | . 5318 , | . 6106, | . 5556 , | . 4882 , | . 5499 , | . 4339, | . 4800, | . 3578 , |
| 0 | FBAR 3-7, |  |  | . 4944 , | . 5017, | .4909, | . 4743 , | . 3900 , | .4642, | . 4375, | . 3882 , | . 3526 , | . 3358 , |
|  | Run title : COD |  |  | F FAROE P | PLATEAU (ICES SUBDIVISION |  |  | Vb1) |  | COD_IND_su |  |  |  |
|  | At | 2/05/2003 |  | 17:09 |  |  |  |  |  |  |  |  |  |
|  |  | Terminal Fs derived using XSA (With F shrinkage) |  |  |  |  |  |  |  |  |  |  |  |
|  |  | Table | 8 | Fishing | mortality | (F) at |  |  |  |  |  |  |  |
|  |  | YEAR, |  | 1973, | 1974, | 1975, | 1976, | 1977, | 1978, | 1979, | 1980, | 1981, | 1982, |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 2, |  |  | . 0657 , | . 0816 , | . 0774 , | .0933, | . 0481 , | . 0588, | .0433, | . 0544 , | . 0523, | . 0585, |
|  | 3, |  |  | . 2322 , | .1568, | . 3193 , | . 1723, | . 3036 , | . 1896, | . 2622, | . 2391, | . 2877, | . 2226, |
|  | 4, |  |  | . 3048 , | . 2046 , | .4359, | . 3665 , | . 4748 , | . 4291, | . 4308 , | . 3695 , | . 3408, | . 3601 , |
|  | 5, |  |  | . 2813 , | . 2953, | . 4134, | . 5568, | . 7532, | . 4289, | . 5049, | . 4337, | . 4368, | . 3886 , |
|  | 6 , |  |  | . 2526 , | . 3797 , | . 4544 , | . 5167, | . 7333, | . 4850, | . 4906 , | . 5181, | . 5643, | . 4046 , |
|  | 7, |  |  | . 3722 , | . 5330, | . 3504 , | .7619, | 1.1138, | . 5968, | . 4480, | . 4119, | . 6939, | . 6925, |
|  | 8, |  |  | . 3259 , | . 3052 , | .4485, | .6429, | . 7776 , | . 5674, | . 6902, | .6437, | . 5014, | . 5525, |
| 9, |  |  |  | . 3091 , | . 3457 , | .4235, | . 5738, | . 7783 , | . 5054, | . 5170, | . 4790 , | . 5115, | .4833, |
| 0 | FBAR 3-7, |  |  | .2886, | . 3139, | . 3947 , | . 4748 , | .6757, | . 4259 , | .4273, | . 3945, | .4647, | .4137, |
|  |  | Table |  | Fishing | mortality |  |  |  |  |  |  |  |  |
|  |  | YEAR, |  | 1983, | 1984, | 1985, | 1986, | 1987, | 1988, | 1989, | 1990, | 1991, | 1992, |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | 2, |  | .0990, | .1071, | . 0655 , | . 0244 , | .0279, | . 0644 , | .1630, | . 0757 , | .0323, | .0199, |
|  |  | 3, |  | . 4669 , | . 3707 , | . 3537 , | . 3523, | . 2173, | . 3365 , | . 4206 , | . 3222 , | . 1963, | .0999, |
|  |  | 4, |  | . 5582, | . 5783 , | . 5065 , | . 6206 , | . 4710 , | . 5506 , | . 7004 , | . 5793, | . 4320, | . 3198, |
|  |  | 5, |  | . 6408, | . 6604, | .6121, | . 7004 , | . 4825, | . 5406 , | . 7275 , | . 6592, | . 5152, | . 3276 , |
|  |  | 6, |  | . 7833, | . 4529, | .9220, | . 8218, | . 5517, | . 7658 , | . 9302, | . 6366, | . 5198, | . 4906 , |
|  |  | 7, |  | 1.0774, | . 4758, | 1.1059, | . 8367, | . 4853, | . 7871 , | 1.0312, | . 7706 , | . 4815, | . 4506 , |
|  |  | 8, |  | . 9412 , | . 4787 , | 1.3183, | . 5384, | . 6172, | . 8492 , | 1.0606, | 1.0308, | . 5954 , | . 3312 , |
|  |  | 9, |  | . 8084, | . 5335, | . 9026 , | . 7103 , | . 5257, | . 7053 , | . 8996 , | . 7425 , | . 5128, | . 3865 , |
| 0 | FBAR | R 3-7, |  | . 7053 , | . 5076, | . 7000 , | . 6664 , | . 4416 , | . 5961 , | . 7620 , | . 5936, | .4289, | . 3377 , |



Table 2.2.6.1.4. Faroe Plateau (Subdivision Vb1) COD. Stock number-at-age.

## Run title : COD FAROE PLATEAU (ICES SUBDIVISION Vb1) At 2/05/2003 17:09 <br> Terminal Fs derived using XSA (With F shrinkage)




Table 2.2.6.1.5. Faroe Plateau (Subdivision Vb1) COD. Summary table.
Run title : COD FAROE PLATEAU (ICES SUBDIVISION Vb1)
COD_IND_su
At 2/05/2003 17:09

Table 16 Summary (without SOP correction)
Terminal Fs derived using XSA (With F shrinkage)

| ', | RECRUITS, <br> Age 2 | TOTALBIO, | TOTSPBIO, | LANDINGS, | YIELD/SSB, | FBAR | 3-7, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961, | 12019, | 65428, | 46439, | 21598, | .4651, |  | . 6059, |
| 1962, | 20654, | 68225, | 43326, | 20967, | .4839, |  | . 5226, |
| 1963, | 20290, | 77602, | 49054, | 22215, | . 4529, |  | . 4944 , |
| 1964, | 21834, | 84666, | 55362, | 21078, | . 3807 , |  | .5017, |
| 1965, | 8269, | 75043, | 57057, | 24212, | . 4244 , |  | .4909, |
| 1966, | 18566, | 83919, | 60629, | 20418, | .3368, |  | .4743, |
| 1967, | 23451, | 105289, | 73934, | 23562, | . 3187 , |  | . 3900, |
| 1968, | 17582, | 110433, | 82484, | 29930, | .3629, |  | . 4642 , |
| 1969, | 9325, | 105537, | 83487, | 32371, | . 3877 , |  | . 4375 , |
| 1970, | 8608, | 98398, | 82035, | 24183, | . 2948 , |  | . 3882 , |
| 1971, | 11928, | 78218, | 63308, | 23010, | . 3635 , |  | . 3526 , |
| 1972, | 21320, | 76439, | 57180, | 18727, | . 3275 , |  | . 3358 , |
| 1973, | 12573, | 107683, | 80516, | 22228, | . 2761 , |  | . 2886 , |
| 1974, | 30480, | 136664, | 95831, | 24581, | . 2565 , |  | . 3139, |
| 1975, | 38320, | 149775, | 105677, | 36775, | . 3480 , |  | . 3947 , |
| 1976, | 18575, | 154920, | 116737, | 39799, | . 3409 , |  | . 4748, |
| 1977, | 9995, | 136019, | 111864, | 34927, | . 3122 , |  | .6757, |
| 1978, | 10749, | 94341, | 76610, | 26585, | . 3470 , |  | .4259, |
| 1979, | 14999, | 83773, | 65382, | 23112, | . 3535 , |  | . 4273, |
| 1980, | 23587, | 84545, | 58390, | 20513, | . 3513 , |  | . 3945 , |
| 1981, | 14004, | 86921, | 62067, | 22963, | . 3700 , |  | . 4647 , |
| 1982, | 22140, | 96656, | 64711, | 21489, | . 3321 , |  | . 4137, |
| 1983, | 25186, | 121715, | 76964, | 38133, | . 4955, |  | . 7053 , |
| 1984, | 47832, | 150401, | 94941, | 36979, | . 3895 , |  | . 5076, |
| 1985, | 17404, | 129879, | 83303, | 39484, | . 4740 , |  | . 7000 , |
| 1986, | 9645, | 98966, | 73173, | 34595, | . 4728 , |  | . 6664, |
| 1987, | 10308, | 78583, | 61964, | 21391, | . 3452 , |  | . 4416 , |
| 1988, | 9019, | 67099, | 52574, | 23182, | .4409, |  | . 5961, |
| 1989, | 16433, | 60600, | 39536, | 22068, | .5582, |  | . 7620 , |
| 1990, | 3684, | 39571, | 30376, | 13487, | . 4440 , |  | . 5936, |
| 1991, | 6675, | 30186, | 22500, | 8750, | . 3889 , |  | . 4289, |
| 1992, | 11477, | 37071, | 22023, | 6396, | . 2904 , |  | . 3377 , |
| 1993, | 10228, | 52724, | 34545, | 6107, | .1768, |  | . 2030, |
| 1994, | 25362, | 86563, | 44775, | 9046, | . 2020, |  | . 1818, |
| 1995, | 43777, | 146947, | 55225, | 23045, | .4173, |  | . 3155 , |
| 1996, | 13106, | 145023, | 86464, | 40422, | .4675, |  | . 6878, |
| 1997, | 6504, | 97773, | 82212, | 34304, | .4173, |  | . 7265 , |
| 1998, | 6494, | 68670, | 57673, | 24005, | . 4162, |  | . 5385, |
| 1999, | 14721, | 68776, | 47940, | 19906, | . 4152 , |  | . 5285, |
| 2000, | 21674, | 96651, | 48411, | 22433, | . 4634 , |  | . 3646 , |
| 2001, | 47706, | 135144, | 63060, | 29409, | . 4664 , |  | . 4457, |
| 2002, | 23427, | 137877, | 68587, | 40159, | . 5855, |  | .8512, |
| Arith. Mean | 18094, | 95493, | 65198, | 24965, | . 3860 , |  | .4837, |
| 0 Units, | (Thousands), | (Tonnes), | (Tonnes), | (Tonnes), |  |  |  |

Table 2.2.6.1.6 Faroe Plateau (Subdivision Vb1) COD. Input to the RCT3 program.
FAROE PLATEAU COD: SURVEYS, 0-GROUP, AND P.PROD

| 'Yrclass' | 'VPA' | 'Ogrpsurv' | 'SS1y' | 'SS2y' | 'SP2y' | 'P.PROD' |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1964 | 18566 | -11 | -11 | -11 | -11 | -11 |
| 1965 | 23451 | -11 | -11 | -11 | -11 | -11 |
| 1966 | 17582 | -11 | -11 | -11 | -11 | -11 |
| 1967 | 9325 | -11 | -11 | -11 | -11 | -11 |
| 1968 | 8608 | -11 | -11 | -11 | -11 | -11 |
| 1969 | 11928 | -11 | -11 | -11 | -11 | -11 |
| 1970 | 21320 | -11 | -11 | -11 | -11 | -11 |
| 1971 | 12573 | -11 | -11 | -11 | -11 | -11 |
| 1972 | 30480 | -11 | -11 | -11 | -11 | -11 |
| 1973 | 38320 | -11 | -11 | -11 | -11 | -11 |
| 1974 | 18575 | -11 | -11 | -11 | -11 | -11 |
| 1975 | 9995 | -11 | -11 | -11 | -11 | -11 |
| 1976 | 10749 | -11 | -11 | -11 | -11 | -11 |
| 1977 | 14999 | -11 | -11 | -11 | -11 | -11 |
| 1978 | 23587 | -11 | -11 | -11 | -11 | -11 |
| 1979 | 14004 | -11 | -11 | -11 | -11 | -11 |
| 1980 | 22140 | -11 | -11 | -11 | -11 | -11 |
| 1981 | 25186 | -11 | -11 | -11 | -11 | -11 |
| 1982 | 47832 | -11 | -11 | -11 | -11 | -11 |
| 1983 | 17404 | -11 | -11 | -11 | -11 | -11 |
| 1984 | 9645 | -11 | -11 | -11 | -11 | -11 |
| 1985 | 10308 | -11 | -11 | -11 | -11 | -11 |
| 1986 | 9019 | -11 | -11 | -11 | -11 | -11 |
| 1987 | 16433 | -11 | -11 | -11 | -11 | -11 |
| 1988 | 3684 | -11 | -11 | -11 | -11 | -11 |
| 1989 | 6675 | 78 | -11 | -11 | -11 | 289 |
| 1990 | 11477 | 523 | -11 | -11 | -11 | 553 |
| 1991 | 10228 | 17 | -11 | -11 | -11 | 708 |
| 1992 | 25362 | 120 | -11 | -11 | 524 | 1008 |
| 1993 | 43777 | 1193 | -11 | -11 | 797 | 1182 |
| 1994 | 13106 | 664 | -11 | 830 | 268 | 1229 |
| 1995 | 6504 | 59 | 38 | 566 | 98 | 674 |
| 1996 | 6494 | 380 | 70 | 518 | 52 | 765 |
| 1997 | 14721 | 1196 | -11 | 372 | 654 | 845 |
| 1998 | 21674 | 8676 | 111 | 1344 | 265 | 1079 |
| 1999 | 47706 | 6202 | 440 | 3375 | 1371 | 1624 |
| 2000 | -11 | 2661 | 205 | 2289 | 346 | 1578 |
| 2001 | -11 | 2760 | 697 | -11 | 123 | 355 |
| 2002 | -11 | 1502 | -11 | -11 | -11 | -11 |

Table 2.2.6.1.7
Analysis by RCT3 ver3.1 of data from file :

## COD8RCT3.TXT

FAROE PLATEAU COD: SURVEYS, 0-GROUP, AND P.PROD


Table 2.2.7.1.1 Faroe Plateau (Subdivision Vb1) COD. Input to management option table.

Recruitment

|  | XSA | RCT3 | Geomean <br> $61-01$ |
| :--- | :--- | :--- | :--- |
| YC2000 | 23427 | 25284 |  |
| YC2001 |  | 12126 |  |
| YC2002 |  | 17740 |  |
| YC2003 |  |  | 15307 |

Maturity

| Age | Observed Av.01-03 Av.01-03 |  |  |
| :---: | :---: | :---: | :---: |
|  | 2003 | 2004 | 2005 |
|  | 0 | 0.04 | 0.04 |
|  | 0.29 | 0.38 | 0.38 |
|  | 0.79 | 0.8 | 0.8 |
|  | 0.88 | 0.93 | 0.93 |
|  | 0.98 | 0.97 | 0.97 |
|  | 1 | 0.99 | 0.99 |
|  | 1 | 1 | 1 |
|  | 1 | 1 | 1 |
|  | 1 | 1 | 1 |

Stock size

| 2003 |
| ---: |
| 12126 |
| 18580 |
| 22073 |
| 4450 |
| 1067 |
| 434 |
| 96 |
| 98 |

$=25284^{*} \exp (-0.2-0.1081)$

Exploitation pattern (not rescaled)
F for ages 7 and 8 removed in 1997
Weights

| Est.from |  |  |
| ---: | ---: | ---: |
| sp. survey | As2002 | As2002 |
| 2003 | 2004 | 2005 |
| 1.0175 | 1.0170 | 1.0170 |
| 1.4870 | 1.7680 | 1.7680 |
| 2.1845 | 2.8050 | 2.8050 |
| 3.1063 | 3.5290 | 3.5290 |
| 3.5630 | 4.0950 | 4.0950 |
| 5.4120 | 4.4750 | 4.4750 |
| 5.9629 | 4.6500 | 4.6500 |
| 6.3712 | 6.2440 | 6.2440 |
| 7.4568 | 7.4568 | 7.4568 |

Table 2.2.7.1.2 Faroe Plateau (Subdivision Vb1) COD. Management option table.

MFDP version 1
Run: Run3
Index file 4/5-2002
Time and date: 17:00 08/05/03
Fbar age range: 3-7

| $2003$ <br> Biomass | SSB | FMult | FBar | Landings |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 109356 | 65540 | 1.0000 | 0.4739 | 31807 |  |  |
| 2004 |  |  |  |  | 2005 |  |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 118053 | 81601 | 0.0000 | 0.0000 | 0 | 143397 | 104958 |
| . | 81601 | 0.1000 | 0.0474 | 4213 | 138655 | 100673 |
| . | 81601 | 0.2000 | 0.0948 | 8246 | 134116 | 96579 |
| . | 81601 | 0.3000 | 0.1422 | 12109 | 129771 | 92666 |
| . | 81601 | 0.4000 | 0.1895 | 15808 | 125611 | 88927 |
| . | 81601 | 0.5000 | 0.2369 | 19350 | 121629 | 85353 |
| . | 81601 | 0.6000 | 0.2843 | 22744 | 117815 | 81936 |
| . | 81601 | 0.7000 | 0.3317 | 25995 | 114162 | 78670 |
| . | 81601 | 0.8000 | 0.3791 | 29110 | 110664 | 75547 |
| . | 81601 | 0.9000 | 0.4265 | 32095 | 107312 | 72562 |
| . | 81601 | 1.0000 | 0.4739 | 34956 | 104102 | 69707 |
| . | 81601 | 1.1000 | 0.5213 | 37699 | 101025 | 66977 |
| . | 81601 | 1.2000 | 0.5686 | 40328 | 98076 | 64366 |
| . | 81601 | 1.3000 | 0.6160 | 42849 | 95250 | 61869 |
| . | 81601 | 1.4000 | 0.6634 | 45267 | 92541 | 59480 |
| . | 81601 | 1.5000 | 0.7108 | 47586 | 89944 | 57195 |
| . | 81601 | 1.6000 | 0.7582 | 49810 | 87454 | 55009 |
| . | 81601 | 1.7000 | 0.8056 | 51944 | 85065 | 52918 |
| . | 81601 | 1.8000 | 0.8530 | 53992 | 82775 | 50916 |
| . | 81601 | 1.9000 | 0.9004 | 55957 | 80577 | 49001 |
| . | 81601 | 2.0000 | 0.9477 | 57844 | 78468 | 47168 |

Input units are thousands and kg - output in tonnes

Table 2.2.7.4.1 Faroe Plateau (Subdivision Vb1) COD. Input to yield-per-recruit calculations (long-term prediction).

| Input to Yield per recruit |  |  |
| :--- | :--- | :--- |
|  |  |  |
| Exploitation | Weightatage | PropMature |
| pattern |  |  |
| Average |  |  |
| 1961-2002 | Average | Average |
| Not rescaled | $1961-2002$ | $1983-2003$ |


| Age 2 | 0.0854 | 1.023 | 0.03 |
| :--- | ---: | ---: | ---: |
| Age 3 | 0.2631 | 1.735 | 0.56 |
| Age 4 | 0.4087 | 2.613 | 0.85 |
| Age 5 | 0.4937 | 3.5 | 0.95 |
| Age 6 | 0.567 | 4.357 | 0.99 |
| Age 7 | 0.6858 | 5.401 | 0.99 |
| Age 8 | 0.6568 | 6.484 | 1 |
| Age 9 | 0.5534 | 7.895 | 1 |

Table 2.2.7.4.2 Faroe Plateau (Subdivision Vb1) COD. Output from yield-per-recruit calculations (long-term prediction).

MFYPR version 1
Run: YLD1
Time and date: 18:32 06/05/03

| FMult | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | SpwnNosSpwn | SSBSpwn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.0000 | 0.0000 | 0.0000 | 0.0000 | 4.4029 | 13.9604 | 2.9365 | 11.9448 | 2.9365 | 11.9448 |
| 0.1000 | 0.0484 | 0.1331 | 0.5233 | 4.0854 | 12.2277 | 2.6287 | 10.2390 | 2.6287 | 10.2390 |
| 0.2000 | 0.0967 | 0.2343 | 0.8779 | 3.8210 | 10.8280 | 2.3734 | 8.8646 | 2.3734 | 8.8646 |
| 0.3000 | 0.1451 | 0.3122 | 1.1164 | 3.5987 | 9.6884 | 2.1599 | 7.7488 | 2.1599 | 7.7488 |
| 0.4000 | 0.1935 | 0.3732 | 1.2748 | 3.4103 | 8.7529 | 1.9799 | 6.8356 | 1.9799 | 6.8356 |
| 0.5000 | 0.2418 | 0.4216 | 1.3784 | 3.2490 | 7.9785 | 1.8267 | 6.0823 | 1.8267 | 6.0823 |
| 0.6000 | 0.2902 | 0.4608 | 1.4443 | 3.1099 | 7.3320 | 1.6953 | 5.4558 | 1.6953 | 5.4558 |
| 0.7000 | 0.3386 | 0.4929 | 1.4846 | 2.9889 | 6.7878 | 1.5817 | 4.9306 | 1.5817 | 4.9306 |
| 0.8000 | 0.3869 | 0.5198 | 1.5073 | 2.8828 | 6.3259 | 1.4828 | 4.4867 | 1.4828 | 4.4867 |
| 0.9000 | 0.4353 | 0.5425 | 1.5181 | 2.7891 | 5.9305 | 1.3959 | 4.1086 | 1.3959 | 4.1086 |
| 1.0000 | 0.4837 | 0.5621 | 1.5210 | 2.7056 | 5.5895 | 1.3192 | 3.7841 | 1.3192 | 3.7841 |
| 1.1000 | 0.5320 | 0.5791 | 1.5187 | 2.6309 | 5.2932 | 1.2509 | 3.5034 | 1.2509 | 3.5034 |
| 1.2000 | 0.5804 | 0.5941 | 1.5128 | 2.5636 | 5.0337 | 1.1898 | 3.2589 | 1.1898 | 3.2589 |
| 1.3000 | 0.6288 | 0.6075 | 1.5048 | 2.5026 | 4.8049 | 1.1348 | 3.0446 | 1.1348 | 3.0446 |
| 1.4000 | 0.6771 | 0.6196 | 1.4955 | 2.4470 | 4.6019 | 1.0850 | 2.8555 | 1.0850 | 2.8555 |
| 1.5000 | 0.7255 | 0.6305 | 1.4854 | 2.3961 | 4.4207 | 1.0398 | 2.6875 | 1.0398 | 2.6875 |
| 1.6000 | 0.7739 | 0.6405 | 1.4749 | 2.3493 | 4.2579 | 0.9985 | 2.5376 | 0.9985 | 2.5376 |
| 1.7000 | 0.8222 | 0.6496 | 1.4643 | 2.3060 | 4.1110 | 0.9606 | 2.4030 | 0.9606 | 2.4030 |
| 1.8000 | 0.8706 | 0.6581 | 1.4538 | 2.2659 | 3.9776 | 0.9257 | 2.2816 | 0.9257 | 2.2816 |
| 1.9000 | 0.9190 | 0.6660 | 1.4434 | 2.2286 | 3.8561 | 0.8934 | 2.1716 | 0.8934 | 2.1716 |
| 2.0000 | 0.9673 | 0.6733 | 1.4333 | 2.1938 | 3.7448 | 0.8635 | 2.0715 | 0.8635 | 2.0715 |


| Reference point | F multiplier | Absolute F |
| :--- | :---: | :---: |
| Fbar(3-7) | 1.0000 | 0.4837 |
| FMax | 0.9981 | 0.4827 |
| F0.1 | 0.5579 | 0.2698 |
| F35\%SPR | 0.8797 | 0.4255 |
| $\quad$ Flow | 2.3119 | 1.1182 |
| $\quad$ Fmed | 0.9914 | 0.4795 |
| $\quad$ Fhigh | 0.1221 | 0.059 |

[^1]Faroe Plateau cod


Figure 2.2.2.1 Faroe Plateau (Subdivision VB1) COD. Catch in numbers.

## Commercial landings



Figure 2.2.3.1 Faroe Plateau (Subdivision VB1) COD. Mean weight-at-age 1961-2002.

Faroe Plateau Cod


Figure 2.2.4.1 Faroe Plateau (Subdivision VB1) COD. Proportion mature-at-age as observed in the spring groundfish survey.


Figure 2.2.5.1 Faroe Plateau (Subdivision VB1) COD. Catch per unit effort in the spring, and summer groundfish survey.


Figure 2.2.5.2 Faroe Plateau (Subdivision VB1) COD. Catch curves from the spring groundfish survey.

Faroe Plateau cod


Figure 2.2.5.3 Faroe Plateau (Subdivision VB1) COD. Catch curves from the summer groundfish survey.

Faroe Plateau cod


Figure 2.2.6.1.1 Faroe Plateau (Subdivision VB1) COD. Catch per unit effort for Cuba trawlers and longliners.


Figure 2.2.6.1.2 Faroe Plateau (Subdivision VB1) COD. Catch per unit effort for Cuba trawlers segregated to age classes.


Figure 2.2.6.1.3 Faroe Plateau (Subdivision VB1) COD. Catch per unit effort for longliners segregated to age classes.


## Summer survey

Summer survey

Figure 2.2.6.1.4 Faroe Plateau (Subdivision VB1) COD. Log-catchability residuals (summer survey).

## Retrospective analysis



Retrospective analysis


Figure 2.2.6.1.5 Faroe Plateau (Subdivision VB1) COD. Results from retrospective analysis.

## Retrospective analysis



Retrospective analysis


Figure 2.2.6.1.5 Faroe Plateau (Subdivision VB1) COD. Results from retrospective analysis. Continued.

## Faroe Plateau cod



Figure 2.2.6.1.6 Faroe Plateau (Subdivision VB1) COD. Results from different XSA runs.

## Faroe Plateau cod


$\longrightarrow$ S.survey alone (SPALY)
$\cdots-$ Cuba trwl. alone, -age 2

- $\quad$ Longliners alone, only 30 iterations

Figure 2.2.6.1.6 Faroe Plateau (Subdivision VB1) COD. Results from different XSA runs. Continued.

## Faroe Plateau cod



Figure 2.2.6.1.6 Faroe Plateau (Subdivision VB1) COD. Results from different XSA runs. Continued.

## Faroe Plateau cod



Figure 2.2.6.1.6. Faroe Plateau (Subdivision VB1) COD. Results from different XSA runs. Continued.

Spawning stock and recruitment


Yield and fishing mortality


Figure 2.2.6.1.7 Faroe Plateau (Subdivision VB1) COD. Yield and fishing versus year. Spawning stock biomass (SSB) and recruitment versus year.

Faroe Plateau cod


Figure 2.2.6.1.8 Faroe Plateau (Subdivision VB1) COD. Yield divided by exploitable biomass (ages 3 and older).


Figure 2.2.6.2.1 Faroe Plateau (Subdivision VB1) COD. Spawning stock - recruitment relationship 1961-2000. Years are shown at each data point.


Figure 2.2.7.2.1 Faroe Plateau (Subdivision VB1) COD. Spawning stock biomass versus fishing mortality 1961-2001. Output from standard graph software.


MFYR version 1
Run: YLD1
Run: Run3
ndex file 4/5-2002
Time and date: 17:00 08/05/03
bar age range: 3-7
Input units are thousands and kg - output in tonnes

|  |  |  |
| :--- | :---: | :---: |
| Reference point | F multiplier | Absolute $\mathbf{F}$ |
| Fbar(3-7) | 1.0000 | 0.4837 |
| FMax | 0.9981 | 0.4827 |
| F0.1 | 0.5579 | 0.2668 |
| F35\%SPR | 0.8797 | 0.4255 |
| Flow | 2.3119 | 1.1182 |
| Fmed | 0.914 | 0.4795 |
| Fhigh | 0.1221 | 0.0590 |

Figure 2.2.7.4.1 Faroe Plateau (Subdivision VB1) COD. Yield-per-recruit and spawning stock biomass (SSB) per recruit versus fishing mortality (left figure). Landings and SSB versus Fbar (3-7).

## Faroe Plateau cod



Figure 2.2.9.1 Faroe Plateau (Subdivision VB1) COD. Results from different XSA runs. Note the deviating fishing mortalities in the period 1996-1999.

### 2.3.1 Trends in landings and effort

Total nominal landings of the Faroe Bank cod from 1986 to 2002 as officially reported to ICES are given in Table 2.3.1.1 and since 1965 in Figure 2.3.1.1. Landings have been highly irregular from 1965 to the mid-1980s, reflecting the opportunistic nature of the fishery on the Bank, with peak landings slightly exceeding 5000 t in 1973. The evolution of landings has been smoother since 1987, declining from about 3500 t in 1987 to only 330 t in 1992 before increasing to 3600 t in 1997. In 2002, 1900 t were reported from the Faroe Bank. Most of the Faroese catch has been taken by pair trawlers and longliners (Table 2.3.1.2).

The decreasing trend in the cod landings from the Faroe Bank lead ACFM in 1990 to advise the Faroese authorities to close the Bank to all fishing. This advice was followed for depths shallower than 200 meters. In 1992 and 1993 longliners and jiggers were allowed to participate in an experimental fishery inside the 200 -meter depth contour. For the quota year 1 September 1995 to 31 August 1996 a fixed quota of 1050 t was set. The new management regime with fishing days was introduced on 1 June 1996 allowing longliners and jiggers to fish inside the $200-\mathrm{m}$ contour. The trawlers are allowed to fish outside the $200-\mathrm{m}$ contour.

### 2.3.2 Stock assessment

Biological samples have been taken from commercial landings since 1974 (the 2002 sampling intensity is shown in the text table below) and from the groundfish survey since 1983. In 2000, an attempt was made to assess the stock using XSA with catch-at-age for 1992-1999, using the spring groundfish survey as a tuning series (1995-1999), but the WG and ACFM concluded that it could only be taken as indicative due to scarce catch-at-age data. No attempt was made to update the XSA in subsequent years given the poor sampling for age composition particularly for trawl landings. The Working Group considered it unwise to calculate an indicative XSA that could be misleading given the poor sampling of an important gear sector.

Sampling from commercial fleets in 2002 is as follows :

Table 2.3.2.1. Samples of lengths, otoliths, and individual weights of Faroe Bank cod in 2002.

| Fleet | Size | Samples | Length | Otoliths | Weights |
| :--- | :--- | ---: | ---: | ---: | ---: |
| Longliners | $<100$ GRT | 2 | 432 | 60 | 60 |
| Longliners | $>100$ GRT | 9 | 1,828 | 329 | 240 |
| Jiggers |  | 2 | 403 | 60 | 60 |
| Sing. trawlers | $<400 ~ H P$ | 0 | 0 | 0 | 0 |
| Sing. trawlers | $400-1000 ~ H P$ | 0 | 0 | 0 | 0 |
| Sing. trawlers | $>1000 ~ H P$ | 0 | 0 | 0 | 0 |
| Pair trawlers | $<1000 ~ H P$ | 1 | 142 | 0 | 0 |
| Pair trawlers | $>1000 ~ H P$ | 4 | 809 | 60 | 60 |
| Total | 18 | 3,614 | 509 | 420 |  |

The Faroese groundfish surveys cover the Faroe Bank and cod is mainly taken within the $200-\mathrm{m}$ depth contour. The catches of cod per trawl hour in depths shallower than 200 meter are shown in Figure 2.3.2.1. The CPUE was low during 1988 to 1995 , varying between 246 and 637 kg /tow since 1996. The 2003 value (717.5) is higher than the 2002 one (443.8). Although noisy, the survey suggests a higher, possibly increasing biomass since 1995.

The length distributions in the 1983-2003 surveys illustrated in Figure 2.3.2.2 show substantially higher numbers in 1996-2003 compared to previous years. They also show that the 1996 year class is extremely weak, since no fish in the size range $40-65 \mathrm{~cm}$ in 1998 ( 2 years old) are observed. In 1999, 2001 and in 2003 the proportion of small fish is large compared to other years, indicating good recruitment.

Figure 2.3.2.3 shows a positive correlation between the survey index and the landings in the same year. The ratio of landings to the survey cpue index provides an exploitation ratio (Figure 2.3.2.4), which can be used as a proxy to relative changes in fishing mortality. The results suggest that fishing mortality has decreased over time and is now close to the lowest observed.

A stock-production model was fitted to landings and the cpue from the survey using ASPIC. The software requires starting guesses for $r$, the intrinsic rate of increase, MSY, $\mathrm{B}_{1} / \mathbf{B}_{\text {MSY }}$ ratio and $q$, catchability coefficients. There was insufficient time to verify the stability of ASPIC to different starting guesses of these parameters, but a retrospective analysis was used, reducing the time-series progressively. The parameter estimates from ASPIC were: MSY $=5015 \mathrm{t}$, $\mathbf{B}_{\mathrm{MSY}}=20960 \mathrm{t}, \mathbf{F}_{\mathrm{MSY}}=0.2393$. The 2003 biomass is estimated to be at $1.25 \mathbf{B}_{\mathrm{MSY}}$. ASPIC assumes that fishing effort is the force driving stock dynamics. This is unlikely to be the case for Faroe Bank Cod where hydro-climatic conditions are likely to have greater influence. ASPIC is therefore unlikely to provide useful results for this stock. The results are presented in Table 2.3.2.5 and Figures 2.3.2.6 and 2.3.2.7 for information.

### 2.3.2.1 Comment on the assessment

An XSA was attempted in the 2000 assessment, but not since. The NWWG concludes that there is poor sampling for age composition, particularly for the trawler landings whose catch is not separated into Faroe Bank or Faroe Plateau during the same trips. Therefore, XSA is not considered useful until reliable coverage of the total catch-at-age can be obtained. The Working Group will attempt using a statistical age-structured model next year which would not be as sensitive to having reliable estimates of the catch-at-age data.

### 2.3.3 Reference points

There is no analytical basis to suggest reference points based on XSA or an accepted general production analysis.

### 2.3.4 Management considerations

The landing estimates are uncertain because since 1996 vessels are allowed to fish both on the Plateau and on Faroe Bank during the same trip, rendering landings from both areas uncertain. Given the relative size of the two fisheries, this is a bigger problem for Faroe Bank cod than for Faroe Plateau cod, but the magnitude remains unquantified for both. The ability to provide advice depends on the reliability of input data. If the cod landings from Faroe Bank are not known, it is difficult to provide advice on landings. If the fishery management agency intends to manage the two fisheries to protect the productive capacity of each individual unit, then it is necessary to regulate the catch removed from each stock. Simple measures should make it possible to identify whether the catch originates from the Bank or from the Plateau, e.g. by storing it in different sections of the hold.

Table 2.3.1.1 Faroe Bank (Subdivision Vb2) COD. Nominal catches (tonnes) by countries 1986-2002 as officially reported to ICES. From 1992 the catches by Faroe Islands and Norway are used in the assessment.

|  | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | 1,836 | 3,409 | 2,960 | 1,270 | 289 | 297 | 122 | 264 | 717 | 561 | 2,051 | 3,459 | 3,092 |
| Norway | 6 | 23 | 94 | 128 | 72 | 38 | 32 | 2 | 8 | 40 | 55 | 135 * | 147 * |
| UK (EW/NI) | - | - | - | - | - | - | + | 1 | 1 | - | - ${ }^{2}$ | - ${ }^{2}$ | - ${ }^{2}$ |
| UK (Scotland) ${ }^{1}$ | 63 | 47 | 37 | 14 | 205 | 90 | 176 | 118 | 227 | 551 | 382 | 277 | 265 |
| United Kingdom |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total | 1,905 | 3,479 | 3,091 | 1,412 | 566 | 425 | 330 | 385 | 953 | 1,152 | 2,488 | 3,871 | 3,504 |
| Used in assessment |  |  |  |  | 361 | 335 | 154 | 266 | 725 | 601 | 2,106 | 3,594 | 3,239 |


|  | 1999 | 2000 * | 2001 | 2002 |
| :--- | :---: | :---: | ---: | ---: |
| Faroe Islands | 1,001 |  |  |  |
| Norway | 88 | 49 | 50 * | 25 |
| UK (EN/NI) | $-{ }^{2}$ | 2 | 2 |  |
| UK (Scotland) | 210 | 245 | 288 |  |
| United Kingdom |  | $-{ }^{2}$ |  |  |
| Total | 1,299 | 294 | 338 | 25 |
| Used in assessment | 1,089 | 1,243 | 1,626 | 1,903 |

[^2]Table 2.3.1.2 Faroe Bank (Subdivision Vb2) COD. Landings of Faroese fleets (in percents) of total Faroese catch (gutted weight)

| Year | Open boats |  | ST<400 | Gillnet | Jiggers | ST<1000 | ST>1000 | PT<1000 | PT>1000 | LL>100 | Ind.trwl | Others | Total, gut.w. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.0 | 8.0 | 0.0 | 0.0 | 16.0 | 7.0 | 7.0 | 11.0 | 40.0 | 11.0 | 0.0 | 0.0 | 100 |
| 1993 | 0.0 | 9.3 | 16.9 | 0.0 | 4.6 | 6.3 | 0.0 | 5.5 | 26.6 | 30.4 | 0.0 | 1.3 | 237 |
| 1994 | 0.5 | 8.8 | 31.2 | 2.6 | 5.1 | 8.1 | 6.4 | 2.8 | 20.0 | 12.6 | 1.6 | 0.5 | 645 |
| 1995 | 1.0 | 3.6 | 3.6 | 0.4 | 23.0 | 0.2 | 9.5 | 11.1 | 16.0 | 31.5 | 0.0 | 0.0 | 505 |
| 1996 | 2.3 | 1.2 | 3.2 | 0.1 | 24.3 | 5.0 | 1.6 | 23.9 | 36.7 | 1.5 | 0.0 | 0.1 | 1846 |
| 1997 | 0.4 | 1.9 | 0.4 | 1.5 | 11.4 | 4.5 | 3.4 | 16.9 | 38.4 | 21.2 | 0.0 | 0.0 | 3101 |
| 1998 | 0.1 | 3.8 | 0.5 | 1.3 | 5.7 | 3.1 | 10.1 | 12.8 | 32.4 | 29.8 | 0.3 | 0.0 | 2783 |
| 1999 | 0.4 | 10.5 | 0.1 | 1.7 | 17.9 | 1.8 | 3.0 | 0.1 | 0.9 | 63.6 | 0.0 | 0.1 | 901 |
| 2000 | 0.3 | 5.9 | 0.3 | 0.0 | 1.3 | 0.0 | 9.3 | 17.7 | 51.2 | 14.0 | 0.0 | 0.0 | 1062 |
| 2001 | 4.1 | 9.2 | 2.3 | 0.5 | 4.8 | 2.9 | 9.2 | 12.6 | 26.9 | 27.3 | 0.2 | 0.0 | 1434 |
| 2002 | 10.3 | 3.5 | 0.0 | 0.0 | 0.3 | 0.0 | 1.5 | 5.9 | 33.4 | 45.3 | 0.0 | 0.0 | 1442 |

```
Faroe Bank Cod RV
Page 1
Apr 2003 at 11:33.54
ASPIC -- A Surplus-Production Model Including Covariates (Ver. 3.82)
FIT Mode
Author: Michael H. Prager; NOAA/NMFS/S.E. Fisheries Science Center
ASPIC User's Manual
    1 0 1 ~ P i v e r s ~ I s l a n d ~ R o a d ; ~ B e a u f o r t , ~ N o r t h ~ C a r o l i n a ~ 2 8 5 1 6 ~ U S A ~
is available gratis
from the author.
Ref: Prager, M. H. 1994. A suite of extensions to a nonequilibrium
    surplus-production model. Fishery Bulletin 92: 374-389.
```

CONTROL PARAMETERS USED (FROM INPUT FILE)

| Number of years analyzed: | 38 | Number of bootstrap trials: | 0 |
| :---: | :---: | :---: | :---: |
| Number of data series: | 1 | Lower bound on MSY: | $5.000 \mathrm{E}+02$ |
| Objective function computed: | in effort | Upper bound on MSY: | $1.000 \mathrm{E}+09$ |
| Relative conv. criterion (simplex): | $1.000 \mathrm{E}-08$ | Lower bound on $r$ : | $7.000 \mathrm{E}-02$ |
| Relative conv. criterion (restart): | $3.000 \mathrm{E}-08$ | Upper bound on $r$ : | $2.500 \mathrm{E}+00$ |
| Relative conv. criterion (effort) : | $1.000 \mathrm{E}-04$ | Random number seed: | 2010417 |
| Maximum F allowed in fitting: | 8.000 | Monte Carlo search mode, trials: | 110000 |

PROGRAM STATUS INFORMATION (NON-BOOTSTRAPPED ANALYSIS)
code 0
Normal convergence.

GOODNESS-OF-FIT AND WEIGHTING FOR NON-BOOTSTRAPPED ANALYSIS
R-squared
Loss component number and title
in CPUE

## MANAGEMENT PARAMETER ESTIMATES (NON-BOOTSTRAPPED)

| Parameter |  | Estimate | Formula | Related quantity |
| :---: | :---: | :---: | :---: | :---: |
| MSY | Maximum sustainable yield | $5.015 \mathrm{E}+03$ | $\mathrm{Kr} / 4$ |  |
| K | Maximum stock biomass | $4.191 \mathrm{E}+04$ |  |  |
| $\mathrm{B}_{\text {MSY }}$ | Stock biomass at MSY | $2.096 \mathrm{E}+04$ | K/2 |  |
| $\mathrm{F}_{\text {MSY }}$ | Fishing mortality at MSY | 2.393E-01 | r/2 |  |
| F(0.1) | Management benchmark | $2.154 \mathrm{E}-01$ | $0.9 * \mathbf{F}_{\text {MSY }}$ |  |
| Y(0.1) | Equilibrium yield at F(0.1) | $4.965 \mathrm{E}+03$ | $0.99 * \mathrm{MSY}$ |  |
| B-ratio | Ratio of $\mathrm{B}(2003)$ to $\mathbf{B}_{\text {MSY }}$ | $1.250 \mathrm{E}+00$ |  |  |
| F-ratio | Ratio of $\mathrm{F}(2002)$ to $\mathbf{F}_{\text {MSY }}$ | $3.210 \mathrm{E}-01$ |  |  |
| F01-mult | Ratio of $\mathrm{F}(0.1)$ to $\mathrm{F}(2002)$ | $2.804 \mathrm{E}+00$ |  |  |
| Y-ratio | Proportion of MSY avail in 2003 | $9.374 \mathrm{E}-01$ | $2 * B r-r^{\wedge} 2$ | Ye (2003) $=4.701 \mathrm{E}+03$ |
| Fishing effort at MSY in units of each fishery: |  |  |  |  |
| $\mathbf{F}_{\text {MSY }}(1)$ | Survey CPUE | $1.250 \mathrm{E}+01$ | r/2q( 1) | $\mathrm{f}(0.1)=1.125 \mathrm{E}+01$ |

## Table 2.3.2.5 (Cont'd)

Faroe Bank Cod RV
RESULTS FOR DATA SERIES \# 1 (NON-BOOTSTRAPPED) Survey CPUE

Data type CC: CPUE-catch series

| Obs | Year | Observed CPUE | Estimated CPUE | $\begin{array}{r} \text { Estim } \\ \mathrm{F} \end{array}$ | Observed yield | Model yield | Resid in log scale | Resid in yield |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1965 | * | 1.102E+02 | 0.4067 | $2.341 \mathrm{E}+03$ | $2.341 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 2 | 1966 | * | $1.160 \mathrm{E}+02$ | 0.3151 | $1.909 \mathrm{E}+03$ | $1.909 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 3 | 1967 | * | $1.328 \mathrm{E}+02$ | 0.2262 | $1.569 \mathrm{E}+03$ | $1.569 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 4 | 1968 | * | $1.337 \mathrm{E}+02$ | 0.5546 | $3.871 \mathrm{E}+03$ | $3.871 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 5 | 1969 | * | $1.255 \mathrm{E}+02$ | 0.3750 | $2.457 \mathrm{E}+03$ | $2.457 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 6 | 1970 | * | $1.235 \mathrm{E}+02$ | 0.4657 | $3.002 \mathrm{E}+03$ | $3.002 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 7 | 1971 | * | $1.252 \mathrm{E}+02$ | 0.3181 | $2.079 \mathrm{E}+03$ | $2.079 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 8 | 1972 | * | $1.369 \mathrm{E}+02$ | 0.3031 | $2.168 \mathrm{E}+03$ | $2.168 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 9 | 1973 | * | $1.166 \mathrm{E}+02$ | 0.8318 | $5.067 \mathrm{E}+03$ | $5.067 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 10 | 1974 | * | $9.402 \mathrm{E}+01$ | 0.4212 | $2.068 \mathrm{E}+03$ | $2.068 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 11 | 1975 | * | $9.456 \mathrm{E}+01$ | 0.4123 | $2.036 \mathrm{E}+03$ | $2.036 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 12 | 1976 | * | $9.306 \mathrm{E}+01$ | 0.4646 | $2.258 \mathrm{E}+03$ | $2.258 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 13 | 1977 | * | $1.031 \mathrm{E}+02$ | 0.1781 | $9.590 \mathrm{E}+02$ | $9.590 \mathrm{E}+02$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 14 | 1978 | * | $9.106 \mathrm{E}+01$ | 0.9209 | $4.379 \mathrm{E}+03$ | $4.379 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 15 | 1979 | * | $7.401 \mathrm{E}+01$ | 0.3379 | $1.306 \mathrm{E}+03$ | $1.306 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 16 | 1980 | * | $8.398 \mathrm{E}+01$ | 0.2743 | $1.203 \mathrm{E}+03$ | $1.203 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 17 | 1981 | * | $9.945 \mathrm{E}+01$ | 0.2366 | $1.229 \mathrm{E}+03$ | 1.229E+03 | 0.00000 | $0.000 \mathrm{E}+00$ |
| 18 | 1982 | * | $1.087 \mathrm{E}+02$ | 0.4144 | $2.352 \mathrm{E}+03$ | $2.352 \mathrm{E}+03$ | 0.00000 | $0.000 \mathrm{E}+00$ |
| 19 | 1983 | $7.899 \mathrm{E}+01$ | $1.084 \mathrm{E}+02$ | 0.4180 | $2.367 \mathrm{E}+03$ | $2.367 \mathrm{E}+03$ | 0.31680 | $0.000 \mathrm{E}+00$ |
| 20 | 1984 | $1.752 \mathrm{E}+02$ | $1.097 \mathrm{E}+02$ | 0.3869 | $2.216 \mathrm{E}+03$ | $2.216 \mathrm{E}+03$ | -0.46849 | $0.000 \mathrm{E}+00$ |
| 21 | 1985 | $1.735 \mathrm{E}+02$ | $1.043 \mathrm{E}+02$ | 0.5439 | $2.961 \mathrm{E}+03$ | $2.961 \mathrm{E}+03$ | -0.50913 | $0.000 \mathrm{E}+00$ |
| 22 | 1986 | $2.661 \mathrm{E}+02$ | $1.007 \mathrm{E}+02$ | 0.3624 | $1.905 \mathrm{E}+03$ | $1.905 \mathrm{E}+03$ | -0.97220 | $0.000 \mathrm{E}+00$ |
| 23 | 1987 | $1.640 \mathrm{E}+02$ | $8.796 \mathrm{E}+01$ | 0.7574 | $3.479 \mathrm{E}+03$ | $3.479 \mathrm{E}+03$ | -0.62326 | $0.000 \mathrm{E}+00$ |
| 24 | 1988 | $7.311 \mathrm{E}+01$ | $5.505 \mathrm{E}+01$ | 1.0753 | $3.091 \mathrm{E}+03$ | $3.091 \mathrm{E}+03$ | -0.28386 | $0.000 \mathrm{E}+00$ |
| 25 | 1989 | $3.655 \mathrm{E}+01$ | $3.336 \mathrm{E}+01$ | 0.8105 | $1.412 \mathrm{E}+03$ | $1.412 \mathrm{E}+03$ | -0.09142 | $0.000 \mathrm{E}+00$ |
| 26 | 1990 | $2.324 \mathrm{E}+01$ | $3.150 \mathrm{E}+01$ | 0.2195 | $3.610 \mathrm{E}+02$ | $3.610 \mathrm{E}+02$ | 0.30391 | $0.000 \mathrm{E}+00$ |
| 27 | 1991 | $5.097 \mathrm{E}+01$ | $4.131 \mathrm{E}+01$ | 0.1553 | $3.350 \mathrm{E}+02$ | $3.350 \mathrm{E}+02$ | -0.21029 | $0.000 \mathrm{E}+00$ |
| 28 | 1992 | $2.843 \mathrm{E}+01$ | $5.858 \mathrm{E}+01$ | 0.0503 | $1.540 \mathrm{E}+02$ | $1.540 \mathrm{E}+02$ | 0.72278 | $0.000 \mathrm{E}+00$ |
| 29 | 1993 | $2.576 \mathrm{E}+01$ | $8.569 \mathrm{E}+01$ | 0.0594 | $2.660 \mathrm{E}+02$ | $2.660 \mathrm{E}+02$ | 1.20200 | $0.000 \mathrm{E}+00$ |
| 30 | 1994 | $4.468 \mathrm{E}+01$ | $1.189 \mathrm{E}+02$ | 0.1168 | $7.250 \mathrm{E}+02$ | $7.250 \mathrm{E}+02$ | 0.97881 | $0.000 \mathrm{E}+00$ |
| 31 | 1995 | $9.532 \mathrm{E}+01$ | $1.609 \mathrm{E}+02$ | 0.0715 | $6.010 \mathrm{E}+02$ | $6.010 \mathrm{E}+02$ | 0.52339 | $0.000 \mathrm{E}+00$ |
| 32 | 1996 | $3.803 \mathrm{E}+02$ | $2.027 \mathrm{E}+02$ | 0.1989 | $2.106 \mathrm{E}+03$ | $2.106 \mathrm{E}+03$ | -0.62908 | $0.000 \mathrm{E}+00$ |
| 33 | 1997 | $5.164 \mathrm{E}+02$ | $2.231 \mathrm{E}+02$ | 0.3085 | $3.594 \mathrm{E}+03$ | $3.594 \mathrm{E}+03$ | -0.83925 | $0.000 \mathrm{E}+00$ |
| 34 | 1998 | $6.377 \mathrm{E}+02$ | $2.361 \mathrm{E}+02$ | 0.2627 | $3.239 \mathrm{E}+03$ | $3.239 \mathrm{E}+03$ | -0.99355 | $0.000 \mathrm{E}+00$ |
| 35 | 1999 | $3.685 \mathrm{E}+02$ | $2.774 \mathrm{E}+02$ | 0.0752 | $1.089 \mathrm{E}+03$ | $1.089 \mathrm{E}+03$ | -0.28410 | $0.000 \mathrm{E}+00$ |
| 36 | 2000 | $2.465 \mathrm{E}+02$ | $3.458 \mathrm{E}+02$ | 0.0688 | $1.243 \mathrm{E}+03$ | $1.243 \mathrm{E}+03$ | 0.33839 | $0.000 \mathrm{E}+00$ |
| 37 | 2001 | $5.373 \mathrm{E}+02$ | $4.136 \mathrm{E}+02$ | 0.0753 | $1.626 \mathrm{E}+03$ | $1.626 \mathrm{E}+03$ | -0.26154 | $0.000 \mathrm{E}+00$ |
| 38 | 2002 | $4.439 \mathrm{E}+02$ | $4.744 \mathrm{E}+02$ | 0.0768 | $1.903 \mathrm{E}+03$ | $1.903 \mathrm{E}+03$ | 0.06641 | $0.000 \mathrm{E}+00$ |

* Asterisk indicates missing value(s).

Table 2.3.2.5 (Cont'd)

Faroe Bank Cod RV
UNWEIGHTED LOG RESIDUAL PLOT FOR DATA SERIES \# 1


Observed (O) and Estimated (*) CPUE for Data Series \# 1 -- Survey CPUE


Figure ?????


Figure 2.3.1.1 Faroe Bank (Subdivision Vb2) COD. Reported landings 1965-2002. Since 1992 only catches from Faroese and Norwegian vessels are considered to be taken on Faroe Bank.

Faroe Bank cod


Figure 2.3.2.1 Faroe Bank (Subdivision Vb2) COD. Catch per unit of effort in the spring groundfish survey (up to 2003) and autumn groundfish survey (up to 2002). If one large haul ( 14 tonnes) is replaced by 4 tonnes (more typical for that particular station) the CPUE drops from about $1300 \mathrm{~kg} / \mathrm{hour}$ to about $530 \mathrm{~kg} /$ hour.

Figure 2.3.2.2.Faroe Bank (Subdivision Vb2) COD. Length distributions in the spring survey 1983-2003.



Figure2.3.2.2 (Continued)

## Faroe Bank cod



Figure 2.3.2.3 Faroe Bank (Subdivision Vb2) COD. CPUE in spring survey and landings (1983-2002)

## Faroe Bank cod



Figure 2.3.2.4 Faroe Bank (Subdivision Vb2) COD. Exploitation ratio (ratio of landings to spring survey interpreted as an index of exploitation rate.)


Figure 2.3.2.6 Retrospective analysis of F and CPUE estimates from ASPIC model

## F/Fmsy



B/Bmsy


Figure 2.3.2.7 Retrospective analysis of $\mathrm{F} / \mathbf{F}_{\mathrm{MSY}}$ and $\mathrm{B} / \mathbf{B}_{\mathrm{MSY}}$ from ASPIC model

### 2.4.1

Introduction

Haddock in Faroese Waters, i.e. ICES Subdivisions Vb 1 and Vb 2 and in the southern part of ICES Division IIa, close to the border of Subdivision Vb 1 , are generally believed to belong to the same stock and are treated as one management unit named Faroe haddock. Haddock is distributed all over the Faroe Plateau and the Faroe Bank from shallow water down to more than 450 m . Spawning takes place from late March to the beginning of May with a peak in the middle of April and occurs in several areas on the Faroe Plateau and on the Faroe Bank. The haddock does not form as dense spawning aggregations like cod and saithe, nor does it perform ordinary spawning migrations. After spawning, eggs and fry are pelagic for about 4 months over the Plateau and Bank and settling starts in August. This is a prolonged process, and pelagic juveniles can be found at least until September. During their first years of life some individuals are to be found in pelagic waters and this vertical distribution seems to be connected to year-class strength. No special nursery areas can be found, because young haddock are distributed all over the Plateau and Bank. After settling the haddock is regarded very stationary as seen in tagging experiments. Different growth in different parts of the distribution area as well as a large degree of heterogeneity in genetic investigations support this.

### 2.4.2 Trends in landings and fisheries

Nominal landings of Faroe haddock have in recent years increased very rapidly from only 4000 t in 1993 to more than 25500 t in 2002. About 1-2 000 t have in recent years been taken from Subdivision Vb2, the rest from the Faroe Plateau (Tables 2.4.1 and 2.4.2). As can be seen from Figure 2.4.1, landings in 2002 were at historical high levels and from the cumulative landings by month (Figure 2.4.2) they are expected to increase even further in 2003.

Faroese vessels have taken almost the entire catch in recent years (Figure 2.4.1). Table 2.4.3 shows the Faroese landings since 1985 and the proportion taken by each fleet category. The longliners have been taken most of the catches in recent years followed by the pair trawlers. The last 2-3 years the otterboards trawlers above 1000 HP have been given individual quotas of cod and haddock and consequently their share of the landings has increased.

The 2002 monthly Faroese landings of haddock by fleet category from Subdivisions Vb 1 and Vb 2 , are shown in Figure 2.4.3. The landings from the Plateau were high in the first month of the year until the end of the spawning time in April/May, stayed low during the summer, and increased again in late autumn. On the Faroe Bank the monthly landings were high during the first 4 months and relatively small the rest of the year.

### 2.4.3 Catch-at-age

For the Faroese landings, catch-at-age data were provided for fish taken from the Faroe Plateau and the Faroe Bank. The sampling intensity in 2002 is shown in the table below. Compared to 2001, the number of samples and of individual length measurements was lower in 2002, whereas the number of age readings and of individual weightings of the fish was somewhat higher.

Sampling of Faroese haddock landings from Vb in 2002

|  | Open Boats | $\begin{array}{\|c\|} \hline \text { LLiners } \\ <\text { 100GRT } \\ \hline \end{array}$ | $\begin{array}{\|c\|} \hline \text { LLiners } \\ \gg 100 G R T \end{array}$ | $\begin{array}{\|l\|} \hline \text { OB. trawl. } \\ <400 \mathrm{HP} \end{array}$ | $\begin{array}{\|c\|} \hline \text { OB. trawl. } \\ 400-999 H P \end{array}$ | $\begin{gathered} \text { OB. trawl. } \\ >1000 \mathrm{HP} \end{gathered}$ | Pair trawl. < 1000HP | Pair trawl. $>1000 \mathrm{HP}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No. of samples | 23 | 74 | 81 | 8 | 24 | 5 | 10 | 52 | 277 |
| No. of length measurements | 4525 | 14910 | 17113 | 1667 | 5009 | 1210 | 2527 | 11416 | 58377 |
| No. of aged fish | 480 | 1482 | 2647 | 120 | 603 | 180 | 120 | 1140 | 6772 |
| No. of weighted fish | 360 | 420 | 1260 | 60 | 180 | 60 | 120 | 960 | 3420 |

Samples from each fleet category were disaggregated by season and then raised by the catch proportions to give the 2002 catch-at-age in numbers for each fleet (Table 2.4.4). Catches of some minor fleets have been included under the "Others" heading. No catch-at-age data were available from other nations fishing in Faroese waters. Therefore, catches by UK and France trawlers were assumed to have the same age composition as Faroese otter board trawlers larger than 1000 HP. The Norwegian longliners were assumed to have the same age distribution as the Faroese longliners greater than 100 GRT. The most recent data were revised according to the final catch figures. The resulting total catch-at-age in numbers is given in Tables 2.4.4 and 2.4.5, and in Figure 2.4.4 the LN (catch-at-age in numbers) is shown for the whole period of analytical assessments.

In general the catch-at-age matrix in recent years appears consistent, except for the behaviour of a few small year classes, both in numbers and mean weights-at-age. Also there are some problems with what ages should be included in the plus group; there are some periods where no or only a few fishes are older than 9 years, and other periods with a quite substantial plus group ( $10+$ ). These problems have been adressed in former reports of this WG and will not be further dealt with here.

No estimates of discards of haddock are available. However, since almost no quotas are used in the management of this stock the incitament to discard in order to high grade the catches should be low. The practice to hang up young fish to dry for local consumption as explained for cod (Section 2.2), also applies to haddock but to a much less degree. Moreover there is a ban on discardings. The landings statistics is therefore regarded as being adequate for assessment purposes.

### 2.4.4 Weight-at-age

Mean weight-at-age data are provided for the Faroese fishery (Table 2.4.6). Figure 2.4.5 shows the mean weights-at-age in the landings for age groups $2-7$ since 1976. After an increase for all ages in a few years the weights have been historical high, but are now levelling off and even decreasing for some ages. The same is seen for 2003 when comparing samples from the commercial catches during the first months of the year (Figure 2.4.6). The mean weights-at-age in the catch were also assumed for the stock.

### 2.4.5 Maturity-at-age

Maturity-at-age data is available from the Faroese Spring Groundfish Surveys 1982-2003. The survey is carried out in February-March, so the maturity-at-age is determined just prior to the spawning of haddock in Faroese waters and the determinations of the different maturity stages should be relatively easy. However, when revising the spring survey data, some inconsistencies were detected. These have been corrected in the revised spring survey series back to 1994, as explained for cod (Section 2.2) and in WD14; there could however be some problems further back in time.

In order to reduce eventual year-to-year effects due to possible inadequate sampling and at the same time allow for trends in the series, a 3-year running average has been used by the WG in the assessment. For the years prior to 1982, average maturity-at-age from the surveys 1982-1995 was adopted (Table 2.4.7 and Figure 2.4.7).

### 2.4.6 Assessment

### 2.4.6.1 Tuning and estimates of fishing mortality

Commercial cpue series. Although several commercial catch per unit effort series are updated every year, only two commercial series are used in the evaluations of stock size and fishing mortality. The two commercial series consist of a longliner series including the logbook data from 5 selected longliners larger than 100 GRT (directed effort measured as number of hooks) and a trawler series including logbook data (catch-at-age in numbers and corresponding effort in number of trawl hours) from a homogeneous group of pair trawlers larger than 1000 HP (CUBA), which have been engaged in a mixed saithe, cod, and haddock fishery since the middle of the 1980s.

Fisheries-independent cpue series. Two annual groundfish surveys are available, one carried out in February-March since 1982 ( 100 stations per year down to $500-\mathrm{m}$ depth), and the other in August-September since 1996 ( 200 stations per year down to $500-\mathrm{m}$ depth). Biomass estimates ( $\mathrm{kg} / \mathrm{hour}$ ) are available for both series, but due to problems with the database (see last year's report), age-disaggregated data were only available last year for the summer series. A major revision of the data is ongoing. As last year the whole summer survey series is available, and this year the spring survey was available back to 1994.

Choice of tuning series. In the tuning of this stock it has been standard to combine all available series in one tuning file. Recently it has been brought up that this is not a wise thing to do, and in general the use of commercial CPUE series has been questioned. Therefore the WG last year decided to make several runs using different combinations of series and with each series separately. Also different shrinkages were applied. Retrospective runs for many of these exercises were shown, and ranges of results given. The WG concluded that since the main differences between the runs are the recruitment estimates, it was appropriate just to present detailed XSA diagnostics and VPA results from the run with the summer survey only, using catchability independent of stock size for all ages (prior to this catchability was assumed to be dependent on stock size for ages younger than 3) and a shrinkage of 2.0 as has been normal for this stock in recent years (except in 2000 when shrinkage was set at 0.5 ). This was questioned by ACFM, arguing that this series is very
short and has little information on older ages; in such a case the very light shrinkage also was dangerous. Inclusion of the two commercial series was advised.

This year the WG did some more analysis on the two commercial cpue series as well as on the two survey series. And again it was decided not to use the commercial series for tuning of the presented XSA. Arguments for this are given below.

Retrospective plots with the two commercial series were given in last year's report (and in former reports) and did in general not behave very good. $\log \mathrm{q}$ residuals for the pair trawlers (Cuba), as shown in Figure 2.4.8A, are noisy and with trends, even when removing some ages as has been the common practice in the past (Figure 2.4.8B). The same applies to the longliners (Figure 2.4.9A), and although most of the trends have been removed by shortening the timeseries and removing of some ages (common practice for this series) the series is still very noisy (Figure 2.4.9B).

LN(numbers-at-age) for the summer survey is presented in Figure 2.4.10 and shows a consistent pattern. Also $\log \mathrm{q}$ residuals behave satisfactorily for ages 2-8 (Figure 2.4.11) and the same applies for the relationship between the indices for one age compared with the indices for the same age one year after (Figure 2.4.12). The same analysis on the spring survey is presented in Figures 2.4.11, 2.4.13 and 2.4.15, and with the same conclusions except for the log q residuals for some ages which consequently have been left out of the further analysis. In general there is a good consistency between the indices at age as estimated in the two series (Figure 2.4.14).

Plots of age-aggregated CPUE's (kg/hour) for the two commercial series and for the two surveys are given in Figure 2.4.16. In general, all series show that the stock has increased from a low level in the mid-1990s and after a decrease for a few years have increased again. However, there are apparently som time lags involved, and in a few cases the series give contradictionary signals; see last year's report for more details.

Last year the summer survey alone was used for tuning, with tapered time weighting, catchability independent of stock size for all ages and F shrinkage of 2.0; a retrospective plot of this spaly run this year is shown in Figure 2.4.17. Using tapered time weighting was a mistake, but omitting this gave only marginal differences (not shown but available in WG folder). As explained above the WG this year decided to use the two surveys combined for tuning as they now have more years included and seem to behave reasonably for the most important age groups in the landings. First the same ages as in the spaly run with the same settings (no tapered time weighting) were used and the restrospective plots are shown in Figure 2.4.17. In order to use the results from the XSA for estimation of future recruitment, ages 0 and 1 were included. The results were very similar but for some unknown reason several statistical diagnostics could not be given in the output (were zero). Exploring this it was seen that this was connected to the shringage applied and only with shrinkages heavier than 0.6 , the outputs were satisfactory. After careful inspection of all diagnostics and retrospective plots (Figure 2.4.17), the WG decided to use a shrinkage of 0.5 . Reasons for using shrinkage of 2.0 in recent years were changes in the fleet behaviour in connection with the introduction of the new management system in 1996 and the sudden changes in the stock at the same time. These arguments may not be valid any more. A comparison of these 4 different runs is given in Figure 2.4.18. Figure 2.4.19 compares 11 runs with each tuning series separately and different combinations thereof.

Results. The indices at age for the present tuning series are shown in Table 2.4.8 and the XSA with diagnostics in Table 2.4.9.

The fishing mortalities from the final XSA run are given in Table 2.4.10 and in Figure 2.4.23B. According to this the fishing mortality showed an overall decline since the early 1960s and it has been estimated to be below or at the natural mortality of 0.2 in several years from the late 1970s. Since 1993 it has been increasing again and in 1998 and 1999 it was estimated above 0.5 , but decreased in 2000 and 2001 to 0.37 and 0.40 , respectively. The estimated point value for $\operatorname{Fbar}(3-7)$ from this year's assessment is 0.45 . The reference ages include a high F-value for the very small 1996 YC at age 6; excluding this age will imply a Fbar of 0.38 . In comparisons all the runs with different fleets and different shrinkages gave fishing mortalities in the range of 0.35 to 0.52 (Figure 2.4.19).

### 2.4.6.2 Stock estimates and recruitment

Compared to former assessments, the 2000 assessment changed the perception of stock size (and fishing mortality) considerably and this year's assessment is consistent with this. The stock size in numbers is given in Table 2.4.11 and a summary of the "VPA" with the biomass estimates is given in Tables 2.4.12 and 2.4.18 and in Figure 2.4.23C,D. According to this assessment, the spawning stock biomass decreased from 67000 t in 1987 to 21000 t in 1994, increased to 75000 t in 1997, but has since decreased to about 42000 t in 2000. In 2001 it increased again to 50000 t and again in 2002 to above 70000 t . For comparison, this year's different runs gave SSB's in the range of 61-87 000 t (Figures 2.4.18 and 2.4.19). The decline in the spawning stock began in the late 1970s due to very poor recruitment in
the years before. The stabilization at relatively high SSB's in the mid-1980s was due to the relatively good 1982 and 1983 year classes, but the decline since then was partly due to poor year classes since the mid-1980s, as well as the pronounced decline in the mean weights-at-age in the stock. The main reason for the very abrupt increase in the spawning stock biomass is the recruitment and growth of the very large 1993 year class and the well-above-average 1994 year class. The most recent increase in the spawning stock is due to new strong year classes entering the fishery. In the past there have been considerable doubts about the sizes of incoming year classes. Due to the lack of reliable recruitment indices it has been normal practise to replace XSA estimates with the geometric mean of a reference period's recruitments at age 2 . With the presence of two survey series and inclusion of indices from them for ages outside the commercial catch-at-age the information on incoming year classes has improved; it was not felt worthwhile to repeat the use of RCT3 estimates as at least the same information is derived directly from the XSA. The 1999 YC is now confirmed being the highest on record at age $2(89$ mio. ) and the YC's from 2000 and 2001 are estimated above average (about 47 mio.), Tables 2.4.12, 2.4.18 and Figure 2.4.23C. The different exploratory XSA runs gave estimates for the 1999 YC at age 2 in the order of $90-125$ mio.

### 2.4.7 Prediction of catch and biomass

### 2.4.7.1 Input data

### 2.4.7.1.1 Short-term prediction

The input data for the short-term predictions are given in Tables 2.4.13-14. All year classes up to 2000 are from the final VPA, the 2001-2002 year classes at age 2 are estimated from the XSA at ages 0 and 1 and applying a natural mortality of 0.2 in a foreward calculation of the numbers using basic VPA equations. The YC 2003 at age 2 in 2005 is estimated as the geometric mean of the 2-year-olds in 1980-2004.

The exploitation pattern used in the prediction was derived from averaging the 2000-2002 fishing mortality matrices from the final VPA without rescaling to the recent values. The high F values for age 8 in 2000 and age 6 in 2002 have been excluded from the averaging, since they are very small and presumably only represent noise in the data and they will have a large impact on the prediction of future biomasses and catches. When excluding these to datapoints from the series, a rescaling of the average exploitation pattern to the 2002 reference F will give almost exactly the same result as the unscaled ones. The same exploitation pattern was used for all three years.

The mean weight-at-age for ages 2-10 in 2003-2005 was calculated using the cohort aproach as described by K. Brander in a WD which was circulated to the ACFM members, i.e. mean weight-at-age $a+1=$ mean weight-at-age $a+$ Growth of the same YC. The weights-at-age in 2002 were used as starting points. By inspecting the weights-at-age 2 for recent years (Figure 2.4.5) and also the 2 first months of the year (Figure 2.4.6), they appear very stable and the value for year 2002 were assumed for all the years. Then the remaining weights-at-age were derived by adding the corresponding geometric mean growth for each cohort (age a to age a +1 ). The mean weight for the + group in 2002 was also applied in 2003-2005. The same weights-at-age were used for the catch and for the stock as was done in the assessment.

The maturity ogive for 2003 is based on samples from the Faroese Groundfish Spring Survey 2002 and the ogives in 2004-2005 are estimated as the average of the smoothed 2001-2003 values.

### 2.4.7.1.2 Long-term Prediction

The input data for the long-term yield and spawning stock biomass (yield-per-recruit calculations) are listed in Table 2.4.16. Mean weights-at-age (stock and catch) are averages for the 1977-2002 period. The maturity ogives are averages for the years 1982-2002. The exploitation pattern was derived from the fishing mortality matrix from the final VPA as average F-values for the long time period (1961 onwards), rescaled to the $2002 \mathrm{~F}_{\text {bar }}$ (age 3-7).

### 2.4.7.2 Biological reference points

The yield- and spawning stock biomass-per-recruit (age 2) based on the long-term data are shown in Table 2.4.17 and Figure 2.4.20. $\mathbf{F}_{\max }$ and $\mathbf{F}_{0.1}$ are indicated here as 0.52 and 0.19 , respectively. From Figure 2.4.22, showing the recruit/spawning stock relationship, and from Table 2.4.17, $\mathbf{F}_{\text {low }}, \mathbf{F}_{\text {med }}$, and $\mathbf{F}_{\text {high }}$ were calculated to be $0.05,0.23$ and 0.83 , respectively.

In previous assessments of this stock the Minimum Biological Acceptable Limit (MBAL) was set at 40000 t because the occurrence of good recruitment is considerably higher when the spawning stock biomass is above this value (Figure 2.4.21) and ACFM established $\mathbf{B}_{\mathrm{lim}}=40000 \mathrm{t}$. In the 1998 assessment, the $\mathbf{B}_{\mathrm{pa}}$ was calculated as the value lying 2
standard deviations above $\mathbf{B}_{\text {lim }}$, that is 65000 t . By examining among other things the SSB-R plot, ACFM instead proposed $\mathbf{B}_{\mathrm{pa}}=55000 \mathrm{t}$. The reference point $\mathbf{F}_{\mathrm{pa}}$ was proposed by ACFM as the $\mathbf{F}_{\text {med }}$ value 0.25. The $\mathbf{F}_{\text {lim }}$ is defined to be two standard deviations above $\mathbf{F}_{\mathrm{pa}}$ and was set by ACFM at 0.40 . The SG on Precautionary Reference Points for Advice on Fishery Management (SGPRP - February 2003) suggested that $\mathbf{B}_{\text {lim }}$ for Faroe haddock could be decreased to 20000 t , considering that two strong year classes have been produced at SSB below $\mathbf{B}_{\text {lim }}$. The Working Group considers it premature to change $\mathbf{B}_{\mathrm{lim}}$ at this time. Of the 5 year classes produced at SSB below $\mathbf{B}_{\mathrm{lim}}$, three were very small, and two very strong. The strong year classes are believed to be due to favourable environmental conditions, and there is no guarantee that similarly good environmental conditions would occur again should the SSB decrease below $\mathbf{B}_{\text {lim }}$.

The $\mathbf{F}_{\text {lim }}$ and $\mathbf{F}_{\mathrm{pa}}$ appear to be rather conservative. The fishing mortality has been above $\mathbf{F}_{\text {lim }}$ during one third of the timeseries ( 14 of the 42 years), while it was above $\mathbf{F}_{\mathrm{pa}}$ almost $70 \%$ of the time ( 29 out 42 years). Clearly, there is not a high probability that the stock will collapse at fishing mortality in the vicinity of $\mathbf{F}_{\text {lim }}$, particularly given the current high stock biomass. The average fishing mortality over the time period, $\mathrm{F}=0.35$ could therefore be considered as a candidate for $\mathbf{F}_{\mathrm{pa}}$, with an associated $\mathbf{F}_{\text {lim }}$, using $\mathbf{F}_{\mathrm{pa}} \mathrm{e}^{1.645}:$ assuming a of about 0.30 , giving $\mathbf{F}_{\text {lim }}=0.55$.

The history of the haddock fishery in relation to the present four reference points can be seen in Figure 2.4.23. In the period 1961-69 the fishing mortality was above $\mathbf{F}_{\text {lim }}$ and the spawning stock biomass was below $\mathbf{B}_{\mathrm{pa}}$. Then the fishing mortality decreased and the stock biomass increased, resulting in the stock/fishery being within or close to biological safe limits in most of the 1970s and 1980s. In 1989 the stock went below $\mathbf{B}_{\mathrm{pa}}$ and continued to decrease below $\mathbf{B}_{\mathrm{lim}}$ in 1991. This decrease in SSB continued until the lowest observed SSB was reached in 1994. The biomass has since increased, mainly due to the outstanding high 1993 year class and the well-above long-term average 1994 year class and has since 1996 been above. The fishing mortality has however been above $\mathbf{F}_{\mathrm{pa}}$ since 1996, and except for the year 2000 even above $\mathbf{F}_{\text {lim }}$.

### 2.4.7.3 Projections of catch and biomass

### 2.4.7.3.1 Short-term prediction

In the light of the performance of the new management system, it is not unrealistic to assume the same fishing mortalities in 2003 as in 2002. The fleet is almost the same and the number of fishing days per fleet was only reduced by $1 \%$ for the fishing year 2002-2003 as compared to the seasons since 1998/1999. The catch in 2003 is then predicted to be about 31000 t (the highest on record), and continuing with this fishing mortality will result in a 2004 catch in the same order of magnitude ( 32000 t ). The SSB will in this case stay stable in 2003 and 2004 (about 96000 t - among the highest observed), and decrease to 85000 t in 2005. The results of the short-term prediction are shown in Table 2.4.15 and in Figure 2.4.21.

### 2.4.8 Medium-term projections

Medium-term projections were made in the 2001 assessment and not repeated here.

### 2.4.9 Management considerations

Stock and yield are highly variable due to fluctuations in recruitment, especially when fishing mortality is high. With the present favourable environmental conditions the stock appears to be able to produce enough recruits to stay above $\mathbf{B}_{\mathrm{pa}}$. It is doubtful, however, whether the current high fishing mortality could be maintained without seriously jeopardising stock productivity, should a relatively long period of low productivity occur. It would therefore be prudent to decrease fishing mortality in order to have a buffer SSB, should productivity decrease. Juvenile and young fish area closures for all gears capable of catching these fishes could result in substantial increases in yield.

### 2.4.10 Comments on the assessment

This year the assessment indices from the commercial fleets were not used for tuning of the VPA and the assessment was tuned with the same summer survey (updated) as last year combined with the revised spring survey. The decision to exclude the commercial series was based on retrospective pattern, statistic diagnostics from the XSA and the fact that these series do not have good information on some ages, especially the youngest. Also the results from the Working Group on Icelandic cod in autumn 2000 and a recent study by Guðmundsson and Jónsson (see last year's report), revealing marked trend in catchabilty in CPUE series from commercial fleets, were taken into account. Indices from commercial fleets are still very valuable even if they are not used directly for tuning of the VPA, and they are as such an important source of information on the state of the stock. They gave the same main message in the assessment as the survey, but the terminal point estimates were different.

Compared to the predicted values last year regarding recruitment, exploitable biomass (age $2+$ ), spawning stock biomass and fishing mortality for 2002, this year's estimate for 2002 of recruitment is $100 \%$ higher, exploitable biomass is $55 \%$ higher, spawning stock biomass is $17 \%$ higher, fishing mortality $18 \%$ higher and landings $21 \%$ higher. The main reason for these discrepancies is the poor estimation of recruitment to this stock in recent years. The use of the summer survey and the revised spring survey series in the tuning of the VPA and in the prediction of future recruitment is believed to make this year's assessment and predictions more reliable. The major reason for the discrepancy in fishing mortality derives from a high F value for age 6 in this year's assessment; this is the very small 1996 year class. This age is in the reference ages, and removing it will give a reference F of 0.39 - almost exactly the same as the predicted value.

Table 2.4.1 Faroe Plateau (Subdivision Vb1) HADDOCK. Nominal catches (tonnes) by countries 1982-2002, as officially reported to ICES , and the total Working Group estimate in Vb.

| Country | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Denmark | - | - | - | - | 1 | 8 | 4 | - | - | - | 4,655 |
| Faroe Islands | 10,319 | 11,898 | 11,418 | 13,597 | 13,359 | 13,954 | 10,867 | 13,506 | 11,106 | 8,074 | 164 |
| France | 2 | 2 | 20 | 23 | 8 | 22 | 14 | - | - | - | - |
| Germany | 1 | + | + | + | 1 | 1 | - | + | + | + |  |
| Norway | 12 | 12 | 10 | 21 | 22 | 13 | 54 | 111 | 94 | 125 | 71 |
| UK (Engl. and Wales) | - | - | - | - | - | 2 | - | - | 7 | - | 54 |
| UK (Scotland) |  | - | - | - | - | - | - | - | - | - |  |
| United Kingdom |  |  |  | - |  |  |  |  |  |  |  |
| Total |  |  |  |  |  |  |  |  |  |  |  |
| Working Group estimate ${ }^{4,5}$ | 11,935 | 11,912 | 11,448 | 13,641 | 13,391 | 14,000 | 10,939 | 13,617 | 11,207 | 8,199 | 4,944 |


| Country | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | $2002{ }^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | 3,622 | 3,675 | 4,549 | 9,152 | 16,585 | 19,135 | 16,643 | $13,620{ }^{8}$ | $14,198{ }^{8}$ | 23,299 ${ }^{8}$ |
| France ${ }^{1}$ | - |  |  |  |  | $2^{2,7}$ | - ${ }^{\text {² }}$ | 6 | $7{ }^{2}$ | 5 |
| Germany | - |  | 5 | - | - |  | 33 | 1 | 2 | 6 |
| Greenland |  |  |  |  |  |  | $30^{6}$ | $22^{6}$ | $0{ }^{6}$ | $4^{6}$ |
| Norway | 28 | 22 | 28 | 45 | $45^{2}$ | $71^{2}$ | $411^{2}$ | $355{ }^{2}$ | $260{ }^{2}$ | 253 |
| UK (Engl. and Wales) | 81 | 31 | 23 | 5 | $22{ }^{1}$ | $30^{1}$ | $59^{7}$ | $19^{7}$ | $4{ }^{7}$ |  |
| UK (Scotland) ${ }^{3}$ | - | - | - | $\ldots$ | ... | ... |  |  |  |  |
| United Kingdom |  |  |  |  |  |  |  |  |  | $204{ }^{7}$ |
| Total | 3,731 | 3,728 | 4,605 | 9,202 | 16,652 | 19,238 | 17,176 | 14,023 | 14,471 | 23,771 |
| Working Group estimate ${ }^{4,5,8}$ | 4,026 | 4,252 | 4,948 | 9,642 | 17,924 | 22,210 | 18,482 | 15,821 | 16,339 | 25584 |

1) Including catches from Sub-division Vb2. Quantity unknown 1989-1991, 1993 and 1995-2001.
2) Preliminary data
3)From 1983 to 1996 catches included in Sub-division Vb 2 .
3) Includes catches from Sub-division Vb 2 and Division IIa in Faroese waters.
5)Includes French and Greenlandic catches from Division Vb , as reported to the Faroese coastal guard service
4) Reported as Division Vb , to the Faroese coastal guard service.
5) Reported as Division Vb.
6) Includes Faroese landings reported to the NWWG by the Faroese Fisheries Laboratory
7) Included in Vb 2

Table 2.4.2 Faroe Bank ( Subdivision Vb2) HADDOCK. Nominal catches (tonnes) by countries, 1982-2002, as officially reported to ICES, and the total Working Group estimate in Vb 2 .

| Country | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Faroe Islands | 1,533 | 967 | 925 | 1,474 | 1,050 | 832 | 1,160 | 659 | 325 | 217 | 338 |
| France $^{1}$ | - | - | - | - | - | - | - | - | - | - | - |
| Norway $^{\text {UK (Engl. and Wales) }}$ | 1 | 2 | 5 | 3 | 10 | 5 | 43 | 16 | 97 | 4 | 23 |
| UK (Scotland) |  |  |  |  |  |  |  |  |  |  |  |


| Country | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | $2002{ }^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | 185 | 353 | 303 | 338 | 1,133 | 2,810 | 1,110 | 1,565 ${ }^{4}$ | 1,655 ${ }^{4}$ | $1784{ }^{4}$ |
| France ${ }^{1}$ | - | - | - | - | - |  |  |  | $1^{2}$ | 1 |
| Norway | 8 | 1 | $1^{2}$ | $40^{2}$ | $4{ }^{2}$ | $60^{2}$ | $3^{2}$ | 48 | $64^{2}$ | 28 |
| UK (Engl. and Wales) | + | + | 1 | ${ }^{1}$ | ${ }^{1}$ |  | 1 | 1 | 1 |  |
| UK (Scotland) ${ }^{3}$ | 102 | 170 | 39 | 62 | $135{ }^{1}$ | 102 | 193 | 185 | 148 |  |
| Total | 295 | 524 | 343 | 440 | 1,272 | 2,972 | 1,306 | 1,798 | 1,868 | 1,813 |
| Working Group estimate 4) |  |  |  |  |  |  |  |  |  |  |

1) Catches included in Sub-division Vb 1 .
2) Provisional data
3)From 1983 to 1996 includes also catches taken in Sub-division Vb 1 (see Table 2.4.1)
3) Includes Faroese landings reported to the NWWG by the Faroese Fisheries Laboratory

Table 2.4.3
Faroe Plateau (Subdivision Vb1) HADDOCK.
Total Faroese landings of haddock from Division Vb and the contribution (\%) by each fleet category (metier).
In the column to the right are the average haddock percentages of the total landings of all species by each fleet category. Total catch in this table may deviate from official landings.

|  | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | Haddock \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Open boats | 7 | 7 | 11 | 2 | 3 | 2 | 3 | 2 | 1 | 1 | 1 | 2 | 2 | 2 | 2 | 1 | 2 | 3 | 18 |
| Longliners < 100GRT | 39 | 39 | 39 | 49 | 58 | 60 | 56 | 46 | 24 | 18 | 23 | 28 | 31 | 30 | 23 | 24 | 29 | 31 | 38 |
| Longliners > 100GRT | 13 | 12 | 13 | 19 | 18 | 18 | 18 | 22 | 25 | 25 | 38 | 36 | 38 | 40 | 40 | 36 | 38 | 34 | 21 |
| Otterboard trawlers < 400HP | 1 | 2 | 2 | 2 | 1 | 1 | 2 | 2 | 8 | 8 | 7 | 6 | 3 | 2 | 2 | 4 | 2 | 2 | 11 |
| Otter board trawlers 400-999HP | 6 | 3 | 5 | 4 | 3 | 3 | 1 | 1 | 3 | 2 | 5 | 7 | 6 | 6 | 5 | 5 | 5 | 4 | 12 |
| Otterboard trawlers > 1000HP | 8 | 5 | 2 | 2 | 2 | 2 | 2 | 1 | 1 | 3 | 2 | 2 | 3 | 3 | 7 | 5 | 5 | 11 | 1 |
| Pairtrawlers < 1000HP | 19 | 20 | 17 | 11 | 7 | 5 | 7 | 11 | 13 | 10 | 8 | 7 | 6 | 5 | 6 | 7 | 6 | 4 | 7 |
| Pairtrawlers > 1000HP | 6 | 10 | 9 | 9 | 6 | 8 | 11 | 14 | 22 | 29 | 16 | 13 | 12 | 12 | 14 | 19 | 12 | 10 | 4 |
| Nets | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Jigging | 1 | 0 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 1 | 2 | 1 |
| Other gears | 0 | 1 | 1 | 2 | 1 | 1 | 1 | 1 | 3 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 |
| Total catch, tonnes gutted | 3570 | 12967 | 829 | 0697 | 866 | 0319 | 7469 | 4103 | 3275 | 3629 | 4371 | 8535 | 890 | 669 | 062 | 881 | 555 | 842 |  |

Table 2.4.4 Faroe Plateau (Subdivision Vb1) HADDOCK. Catch-at-age in number by fleet category.

| Age | $\begin{gathered} \text { Vb1 } \\ \text { Open } \\ \text { Boats } \end{gathered}$ | Vb1 LLiners $<$ 100GRT | Vb1 LLiners $>100 \mathrm{GRT}$ |  | Vb1 OB. trawl. 400-999HP | $\begin{gathered} \hline \text { Vbl } \\ \text { ob. trawl. } \\ >1000 \mathrm{HP} \end{gathered}$ | Vb1 <br> Pair trawl. <br> $<1000 H P$ | vb1 <br> Pair trawl. $>1000 \mathrm{HP}$ | Vb 1 Others | Vb1 <br> All Faroese <br> Fleets |  | Vb2 <br> All Faroese Paintrawlers | Vb2 All Faroese Fleets | Vb Foreign <br> Trawlers | Vb Foreign LLiners | $\begin{gathered} \hline \text { Vb } \\ \text { Total } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | - | 0 |
| 2 | 27 | 713 | 56.3 | 16 | 38 | 28 | 18 | 34 | 57 | 1497 | 8 | 23 | 32 | 2 | 23 | 1555 |
| 3 | 467 | 5308 | 4021 | 198 | 446 | 1363 | 207 | 731 | 50.5 | 13290 | 536 | 312 | 833 | 121 | 163 | 14406 |
| 4 | 41 | 601 | 877 | 42 | 86 | 571 | 117 | 350 | 106 | 2785 | 38 | 45 | 84 | 50 | 36 | 2954 |
| 5 | 17 | 267 | 372 | 19 | 46 | 152 | 52 | 166 | 43 | 1132 | 26 | 43 | 70 | 13 | 15 | 1231 |
| 6 | 2 | 24 | 51 | 2 | 5 | 17 | 4 | 11 | 5 | 121 | 2 | 9 | 11 | 2 | 2 | 136 |
| 7 | 4 | 68 | 60 | 5 | 10 | 15 | 12 | 39 | 8 | 222 | 11 | 9 | 19 | 1 | 2 | 245 |
| 8 | 15 | 223 | 237 | 20 | 52 | 74 | 40 | 131 | 31 | 824 | 13 | 12 | 25 | 6 | 10 | 865 |
| 9 | 20 | 329 | 339 | 28 | 73 | 79 | 38 | 133 | 41 | 1081 | 11 | 11 | 22 | 7 | 14 | 1124 |
| 10 | 0 | 11 | 11 | 0 | 0 | 1 | 1 | 3 | 1 | 26 | 0 | 1 | 1 | 0 | 0 | 29 |
| 11 | 0 | a | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 4 |
| 12 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 2 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total no. | 594 | 7546 | 6536 | 329 | 757 | 2320 | 490 | 1597 | 798 | 20986 | 644 | 465 | 1098 | 203 | 265 | 22551 |
| Catch, t . | 516 | 6760 | 7037 | 379 | 851 | 2471 | 595 | 1926 | 810 | 21347 | 629 | 572 | 1201 | 216 | 295 | 23049 |

Notes: Numbers in 1000'
Catch, gutted weight in tonnes
Others includes netters, jiggers, ather small categaries and catches not atherwise accounted for
LLiners $=$ Longliners OB.trawl. $=$ Otterboard trawlers $\quad$ Pair Trawil. $=$ Pair trawlers

Table 2.4.5
Faroe Haddock. Catch number-at-age.


Table 2.4.6 Faroe Haddock. Catch weights-at-age.


Table 2.4.7 Faroe Haddock. Proportion mature-at-age.
Run title : FAROE HADDOCK (ICES DIVISION Vb)
HAD_IND
At 7/05/2003 11:06


## Table 2.4.8 Faroe Haddock



```
Lowestoft VPA Version 3.1
    7/05/2003 10:35
Extended Survivors Analysis
FAROE HADDOCK (ICES DIVISION Vb) HAD_IND
CPUE data from file D:\Vpa\vpa2003\yc-est\comb-survey1.dat
Catch data for 42 years. 1961 to 2002. Ages 0 to 10.
    Fleet First Last First Last Alpha Beta
    year year l
llumMER SURVEY 
Time-series weights :
    Tapered time weighting not applied
Catchability analysis :
    Catchability dependent on stock size for ages < 2
        Regression type = C
        Minimum of 5 points used for regression
        Survivor estimates shrunk to the population mean for ages < 2
    Catchability independent of age for ages >= 6
Terminal population estimation :
    Survivor estimates shrunk towards the mean F
    of the final 5 years or the 5 oldest ages.
    S.E. of the mean to which the estimates are shrunk = . 500
    Minimum standard error for population
    estimates derived from each fleet = . 300
    Prior weighting not applied
Tuning converged after 30 iterations
1
Regression weights
\(1.0001 .000 \quad 1.000 \quad 1.000 \quad 1.000 \quad 1.000 \quad 1.000 \quad 1.000 \quad 1.000 \quad 1.000\)
```


## Table 2.4.9 (cont.)

| Fishing mortalities |  |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| 0 | .000 | .000 | .000 | .000 | .000 | .000 | .000 | .000 | .000 | .000 |
| 1 | .000 | .000 | .000 | .000 | .000 | .000 | .000 | .000 | .000 | .000 |
| 2 | .071 | .050 | .010 | .009 | .011 | .040 | .013 | .073 | .052 | .037 |
| 3 | .167 | .165 | .107 | .084 | .105 | .212 | .766 | .322 | .227 | .236 |
| 4 | .188 | .260 | .314 | .371 | .241 | .280 | .299 | .299 | .473 | .336 |
| 5 | .197 | .152 | .310 | .420 | .482 | .373 | .432 | .368 | .448 | .455 |
| 6 | .194 | .227 | .190 | .386 | .533 | .644 | .463 | .477 | .461 | .748 |
| 7 | .198 | .233 | .244 | .374 | .570 | 1.322 | .820 | .370 | .399 | .474 |
| 8 | .159 | .244 | .246 | .370 | .386 | 1.088 | 2.145 | .782 | .317 | .576 |
| 9 | .188 | .224 | .262 | .321 | .591 | .885 | .805 | .437 | .374 | .451 |

XSA population numbers (Thousands)

|  |  |  | AGE |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| 1993 | 1.33E+05 | $7.74 \mathrm{E}+03$ | 1.82E+03 | $2.14 \mathrm{E}+03$ | $1.77 \mathrm{E}+03$ | $3.43 \mathrm{E}+03$ | $3.37 \mathrm{E}+03$ | $2.91 \mathrm{E}+03$ | $9.85 \mathrm{E}+02$ | $1.30 \mathrm{E}+03$ |
| 1994 | $5.94 \mathrm{E}+04$ | $1.09 \mathrm{E}+05$ | $6.34 \mathrm{E}+03$ | $1.39 \mathrm{E}+03$ | $1.48 \mathrm{E}+03$ | $1.20 \mathrm{E}+03$ | $2.30 \mathrm{E}+03$ | $2.27 \mathrm{E}+03$ | $1.96 \mathrm{E}+03$ | $6.88 \mathrm{E}+02$ |
| 1995 | $1.12 \mathrm{E}+04$ | $4.86 \mathrm{E}+04$ | 8.90E+04 | $4.94 \mathrm{E}+03$ | $9.64 \mathrm{E}+02$ | $9.37 \mathrm{E}+02$ | $8.42 \mathrm{E}+02$ | $1.50 \mathrm{E}+03$ | $1.47 \mathrm{E}+03$ | $1.26 \mathrm{E}+03$ |
| 1996 | $4.47 \mathrm{E}+03$ | $9.19 \mathrm{E}+03$ | 3.98E+04 | $7.22 \mathrm{E}+04$ | $3.63 \mathrm{E}+03$ | $5.77 \mathrm{E}+02$ | $5.62 \mathrm{E}+02$ | $5.70 \mathrm{E}+02$ | $9.64 \mathrm{E}+02$ | $9.43 \mathrm{E}+02$ |
| 1997 | $2.27 \mathrm{E}+04$ | $3.66 \mathrm{E}+03$ | $7.53 \mathrm{E}+03$ | $3.23 \mathrm{E}+04$ | $5.44 \mathrm{E}+04$ | $2.05 \mathrm{E}+03$ | $3.10 \mathrm{E}+02$ | $3.13 \mathrm{E}+02$ | $3.21 \mathrm{E}+02$ | $5.45 \mathrm{E}+02$ |
| 1998 | $3.43 \mathrm{E}+04$ | $1.86 \mathrm{E}+04$ | $3.00 \mathrm{E}+03$ | $6.09 \mathrm{E}+03$ | $2.38 \mathrm{E}+04$ | $3.50 \mathrm{E}+04$ | $1.04 \mathrm{E}+03$ | $1.49 \mathrm{E}+02$ | $1.45 \mathrm{E}+02$ | $1.79 \mathrm{E}+02$ |
| 1999 | 1.45E+05 | $2.81 \mathrm{E}+04$ | 1.52E+04 | $2.36 \mathrm{E}+03$ | $4.03 \mathrm{E}+03$ | $1.47 \mathrm{E}+04$ | $1.97 \mathrm{E}+04$ | $4.46 \mathrm{E}+02$ | $3.25 \mathrm{E}+01$ | $4.00 \mathrm{E}+01$ |
| 2000 | $7.09 \mathrm{E}+04$ | $1.19 \mathrm{E}+05$ | $2.30 \mathrm{E}+04$ | $1.23 \mathrm{E}+04$ | $8.98 \mathrm{E}+02$ | $2.45 \mathrm{E}+03$ | $7.82 \mathrm{E}+03$ | $1.02 \mathrm{E}+04$ | 1.61E+02 | $3.12 \mathrm{E}+00$ |
| 2001 | $7.07 \mathrm{E}+04$ | $5.81 \mathrm{E}+04$ | $9.75 \mathrm{E}+04$ | $1.75 \mathrm{E}+04$ | $7.30 \mathrm{E}+03$ | $5.45 \mathrm{E}+02$ | $1.39 \mathrm{E}+03$ | $3.98 \mathrm{E}+03$ | $5.74 \mathrm{E}+03$ | $6.02 \mathrm{E}+01$ |
| 2002 | $3.70 \mathrm{E}+04$ | $5.78 \mathrm{E}+04$ | $4.75 \mathrm{E}+04$ | $7.57 \mathrm{E}+04$ | $1.14 \mathrm{E}+04$ | $3.72 \mathrm{E}+03$ | $2.85 \mathrm{E}+02$ | 7.17E+02 | $2.18 \mathrm{E}+03$ | $3.42 \mathrm{E}+03$ |
| Estimated |  | Population |  | abundance |  | at 1st Jan |  | 2003 |  |  |
|  | $0.00 \mathrm{E}+00$ | $3.03 \mathrm{E}+04$ | $4.74 \mathrm{E}+04$ | $3.75 \mathrm{E}+04$ | $4.90 \mathrm{E}+04$ | $6.69 \mathrm{E}+03$ | $1.93 \mathrm{E}+03$ | $1.10 \mathrm{E}+02$ | $3.65 \mathrm{E}+02$ | $1.01 \mathrm{E}+03$ |

Taper weighted geometric mean of the VPA populations:
$\begin{array}{llllllllll}2.83 \mathrm{E}+04 & 2.34 \mathrm{E}+04 & 1.92 \mathrm{E}+04 & 1.46 \mathrm{E}+04 & 9.27 \mathrm{E}+03 & 5.52 \mathrm{E}+03 & 3.30 \mathrm{E}+03 & 1.95 \mathrm{E}+03 & 9.77 \mathrm{E}+02 & 4.36 \mathrm{E}+02\end{array}$
Standard error of the weighted Log(VPA populations) :

| 1.0389 | 1.0439 | 1.0457 | 1.0141 | 1.0012 | 1.0006 | 1.0124 | .9820 | 1.1478 | 1.4763 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Log-catchability residuals.

Fleet : SUMMER SURVEY

| Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | No dat | for | is fl | at | is age |  |  |  |  |  |
| 1 | 99.99 | 99.99 | 99.99 | 1.29 | 3.31 | -4.88 | -5.78 | 2.47 | 2.27 | 1.32 |
| 2 | 99.99 | 99.99 | 99.99 | -. 07 | . 46 | -. 41 | -. 36 | -. 07 | . 23 | . 21 |
| 3 | 99.99 | 99.99 | 99.99 | -. 13 | -. 08 | -. 48 | 1.35 | -. 26 | -. 23 | -. 17 |
| 4 | 99.99 | 99.99 | 99.99 | -. 17 | . 28 | . 07 | -. 39 | . 08 | . 22 | -. 08 |
| 5 | 99.99 | 99.99 | 99.99 | -. 29 | . 09 | . 00 | . 18 | . 22 | -. 25 | . 04 |
| 6 | 99.99 | 99.99 | 99.99 | -. 13 | . 12 | -. 47 | -. 05 | . 21 | -. 02 | . 35 |
| 7 | 99.99 | 99.99 | 99.99 | -. 23 | -. 63 | . 70 | . 04 | -. 05 | . 21 | . 13 |
| 8 | 99.99 | 99.99 | 99.99 | -. 24 | -. 08 | . 38 | . 31 | . 24 | -. 08 | . 13 |

Mean log-catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q | -5.3460 | -5.1158 | -5.4978 | -5.5925 | -5.5075 | -5.5075 | -5.5075 |
| S.E (Log q) | .3215 | .6105 | .2336 | .1989 | .2645 | .4095 | .2530 |

## Table 2.4.9 (cont.)

Regression statistics :
Ages with $q$ dependent on year class strength
Age Slope $t$-value Intercept RSquare No Pts Reg s.e Mean Log q $\begin{array}{llllllll}1 & 4.43 & -2.477 & -16.25 & .09 & 7 & 4.07 & -4.19\end{array}$ Ages with $q$ independent of year class strength and constant w.r.t. time.

| Age | Slope | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | .90 | .975 | 5.79 | .95 | 7 | .29 | -5.35 |
| 3 | 1.36 | -1.490 | 3.43 | .77 | 7 | .76 | -5.12 |
| 4 | .93 | 1.036 | 5.73 | .98 | 7 | .22 | -5.50 |
| 5 | .93 | 1.653 | 5.77 | .99 | 7 | .16 | -5.59 |
| 6 | 1.02 | -.215 | 5.48 | .97 | 7 | .29 | -5.51 |
| 7 | 1.03 | -.210 | 5.45 | .93 | 7 | .46 | -5.48 |
| 8 | 1.09 | -1.935 | 5.35 | .99 | 7 | .21 | -5.41 |

Fleet : SPRING SURVEY

| Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | -.97 | 1.66 | 2.60 | -.84 | -2.21 | -1.92 | -.12 | .88 | .23 | .69 |
| 1 | -.05 | -.55 | .67 | .99 | -.90 | .22 | .00 | -.07 | -.62 | .30 |
| 2 | -.27 | .06 | .11 | .66 | .94 | -1.66 | .37 | -.27 | .41 | -.34 |
| 3 | -.29 | -.13 | -.25 | .43 | .40 | .23 | -.10 | -.64 | .37 | -.02 |
| 4 | -.18 | -.21 | -.03 | .48 | .49 | .30 | .00 | -1.28 | .36 | .08 |
| 5 | -.26 | -.85 | -.23 | .88 | .67 | -.07 | .28 | -.58 | .08 | .08 |
| 6 | .29 | -.14 | -.03 | .20 | -.61 | .07 | .43 | -.05 | .29 | -.44 |
| 7 | No data for this fleet at this age |  |  |  |  |  |  |  |  |  |

Mean log-catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age | 2 |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q | -5.9242 | -5.8135 | -6.1443 | -6.3546 | -6.7754 |
| S.E (Log q) | .7175 | .3520 | .5188 | .5292 | .3321 |

Regression statistics :
Ages with $q$ dependent on year class strength
Age Slope t-value Intercept RSquare No Pts Reg s.e Mean Log q

| 0 | 1.97 | -1.962 | .34 | .34 | 10 | 1.62 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | .93 | .387 | 5.94 | .80 | 10 | .62 |

Ages with $q$ independent of year class strength and constant w.r.t. time.

| Age | Slope | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | .83 | 1.256 | 6.58 | .87 | 10 | .57 | -5.92 |
| 3 | .89 | 1.603 | 6.18 | .97 | 10 | .29 | -5.81 |
| 4 | .80 | 2.621 | 6.61 | .95 | 10 | .32 | -6.14 |
| 5 | 1.03 | -.181 | 6.32 | .86 | 10 | .57 | -6.35 |
| 6 | .86 | 2.439 | 6.85 | .98 | 10 | .23 | -6.78 |

## Table 2.4.9 (cont.)

Terminal year survivor and $F$ summaries :
Age 0 Catchability dependent on age and year class strength
Year class $=2002$

| Fleet | Estimated Survivors | Int s.e | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled Weights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SUMMER SURVEY | 1. | . 000 | . 000 | . 00 | 0 | . 000 | . 000 |
| SPRING SURVEY | 60348 . | 1.712 | . 000 | . 00 | 1 | . 271 | . 000 |
| $P$ shrinkage mean | 23412. | 1.04 |  |  |  | . 729 | . 000 |
| F shrinkage mean | 0. | . 50 |  |  |  | . 000 | . 000 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 30264. | .89 | .81 | 2 | .907 | .000 |

Age 1 Catchability dependent on age and year class strength
Year class $=2001$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| SUMMER SURVEY | 176632. | 4.444 | . 000 | . 00 | 1 | . 014 | . 000 |
| SPRING SURVEY | 63594. | . 623 | . 026 | . 04 | 2 | . 727 | . 000 |
| $P$ shrinkage mean | 19204. | 1.05 |  |  |  | . 258 | . 000 |
| $F$ shrinkage mean | 0. | . 50 |  |  |  | .000 | .000 |

Weighted prediction :

| Survivors | Int | Ext | $N$ | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | S.e | S.e |  | Ratio |  |
| 47359. | .53 | .36 | 4 | .685 | .000 |

Age 2 Catchability constant w.r.t. time and dependent on age
Year class $=2000$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| SUMMER SURVEY | 46891. | . 343 | . 155 | . 45 | 2 | . 497 | . 030 |
| SPRING SURVEY | 25204. | . 473 | . 270 | . 57 | 3 | . 261 | . 054 |
| $F$ shrinkage mean | 36427. | . 50 |  |  |  | . 242 | . 038 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | S.e |  | Ratio |  |
| 37517. | .24 | .15 | 6 | .628 | .037 |

Age 3 Catchability constant w.r.t. time and dependent on age
Year class $=1999$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| SUMMER SURVEY | 56932. | . 304 | . 160 | . 53 | 3 | . 388 | . 206 |
| SPRING SURVEY | 50510. | . 294 | . 091 | . 31 | 4 | . 423 | . 230 |
| F shrinkage mean | 33544. | . 50 |  |  |  | . 189 | . 328 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | ---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 48981. | .20 | .10 | 8 | .527 | .236 |

## Table 2.4.9 (cont.)

Age 4 Catchability constant w.r.t. time and dependent on age
Year class $=1998$


Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | ---: | ---: | ---: | ---: | ---: |
| at end of year | s.e | S.e |  | Ratio |  |
| 6694. | .16 | .09 | 10 | .590 | .336 |

Age 5 Catchability constant w.r.t. time and dependent on age
Year class $=1997$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | S.e | Ratio |  | Weights | F |
| SUMMER SURVEY | 1958. | . 183 | . 121 | . 66 | 5 | . 575 | . 450 |
| SPRING SURVEY | 1787. | . 248 | . 211 | . 85 | 6 | . 267 | . 485 |
| F shrinkage mean | 2113. | . 50 |  |  |  | . 158 | . 423 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | ---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 1934. | .15 | .09 | 12 | .635 | .455 |

Age 6 Catchability constant w.r.t. time and dependent on age
Year class $=1996$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :--- | ---: | :---: | ---: | ---: | ---: | ---: | ---: |
|  | Survivors | S.e | s.e | Ratio | Weights | F |  |
| SUMMER SURVEY | 122. | .167 | .150 | .90 | 6 | .531 | .698 |
| SPRING SURVEY | 69. | .230 | .171 | .74 | 7 | .295 | 1.021 |
| F shrinkage mean |  | 181. | .50 |  |  | .174 |  |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 110. | .14 | .13 | 14 | .946 | .748 |

Age 7 Catchability constant w.r.t. time and age (fixed at the value for age) 6
Year class $=1995$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Survivors | S.e | s.e | Ratio | Weights | F |  |
| SUMMER SURVEY | 381. | .157 | .103 | .65 | 7 | .577 | .459 |
| SPRING SURVEY | 456. | .208 | .179 | .86 | 7 | .265 | .396 |
| F shrinkage mean |  | 217. | .50 |  |  | .158 |  |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 365. | .13 | .10 | 15 | .792 | .474 |

Table 2.4.9 (cont.)
Age 8 Catchability constant w.r.t. time and age (fixed at the value for age) 6
Year class $=1994$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SUMMER SURVEY | Survivors | s.e | s.e | Ratio | Weights | F |  |
| SPRING SURVEY | 1157. | .156 | .030 | .19 | 7 | .650 | .517 |
| F shrinkage mean | 1222. | .208 | .114 | .55 | 7 | .175 | .495 |
| F |  | 490. | .50 |  |  | .175 |  |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | ---: | ---: | ---: | ---: | ---: |
| at end of year | s.e | S.e |  | Ratio |  |
| 1005. | .14 | .10 | 15 | .747 | .576 |

Age 9 Catchability constant w.r.t. time and age (fixed at the value for age) 6
Year class $=1993$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio | Weights | F |  |
| SUMMER SURVEY | 1743. | .161 | .047 | .29 | 6 | .612 | .460 |
| SPRING SURVEY | 2376. | .204 | .124 | .61 | 7 | .185 |  |
| F shrinkage mean |  | 1484. | .50 |  |  | .356 |  |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | S.e |  | Ratio |  |
| 1787. | .15 | .06 | 14 | .418 | .451 |

Table 2.4.10 Faroe Haddock. Fishing mortality (F) at age.
Run title : FAROE HADDOCK (ICES DIVISION Vb) HAD_IND
At 7/05/2003 11:06
Terminal Fs derived using XSA (With F shrinkage)



Table 2.4.11 Faroe Haddock. Stock number-at-age.


Table 2.4.12 Faroe Haddock.
Run title : FAROE HADDOCK (ICES DIVISION Vb) HAD_IND
$\begin{array}{lll}\text { At } & 7 / 05 / 2003 & 11: 06 \\ & \text { Table } 16 \quad \text { Summary } \quad \text { (without SOP correction) }\end{array}$
Terminal Fs derived using XSA (With F shrinkage)

|  | $\begin{gathered} \text { RECRUITS } \\ \text { Age } 0 \end{gathered}$ | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | FBAR | 3-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 70656 | 81164 | 47797 | 20831 | . 4358 |  | . 5624 |
| 1962 | 44919 | 83420 | 51875 | 27151 | . 5234 |  | . 6506 |
| 1963 | 33781 | 80753 | 49547 | 27571 | . 5565 |  | . 7002 |
| 1964 | 30143 | 68577 | 44128 | 19490 | . 4417 |  | . 4753 |
| 1965 | 37827 | 65655 | 45555 | 18479 | . 4056 |  | . 5260 |
| 1966 | 81816 | 60934 | 43953 | 18766 | . 4270 |  | . 5288 |
| 1967 | 47691 | 60206 | 41959 | 13381 | . 3189 |  | . 4030 |
| 1968 | 53082 | 78075 | 45379 | 17852 | . 3934 |  | . 4377 |
| 1969 | 23049 | 83814 | 53422 | 23272 | . 4356 |  | . 4853 |
| 1970 | 49492 | 87297 | 59858 | 21361 | . 3569 |  | . 4762 |
| 1971 | 35342 | 81751 | 62908 | 19393 | . 3083 |  | . 4564 |
| 1972 | 78073 | 83079 | 61975 | 16485 | . 2660 |  | . 3964 |
| 1973 | 104507 | 82753 | 61578 | 17976 | . 2919 |  | . 2894 |
| 1974 | 83496 | 95415 | 64631 | 14773 | . 2286 |  | . 2206 |
| 1975 | 39073 | 121785 | 75405 | 20715 | . 2747 |  | . 1799 |
| 1976 | 52361 | 135610 | 89220 | 26211 | . 2938 |  | . 2475 |
| 1977 | 4154 | 121037 | 96373 | 25555 | . 2652 |  | . 3873 |
| 1978 | 7376 | 120570 | 97226 | 19200 | . 1975 |  | . 2781 |
| 1979 | 5208 | 97683 | 85394 | 12418 | . 1454 |  | . 1551 |
| 1980 | 23623 | 87636 | 81902 | 15016 | . 1833 |  | . 1779 |
| 1981 | 29262 | 78962 | 75846 | 12233 | . 1613 |  | . 1813 |
| 1982 | 60838 | 68306 | 56804 | 11937 | . 2101 |  | . 3308 |
| 1983 | 58811 | 63961 | 51811 | 12894 | . 2489 |  | . 2654 |
| 1984 | 39456 | 83382 | 53820 | 12378 | . 2300 |  | . 2284 |
| 1985 | 14055 | 93973 | 62602 | 15143 | . 2419 |  | . 2761 |
| 1986 | 27946 | 98502 | 65604 | 14477 | . 2207 |  | . 2237 |
| 1987 | 21578 | 87615 | 67294 | 14882 | . 2211 |  | . 2643 |
| 1988 | 13499 | 77373 | 61882 | 12178 | . 1968 |  | . 2010 |
| 1989 | 4367 | 69699 | 51703 | 14325 | . 2771 |  | . 2854 |
| 1990 | 3968 | 53552 | 43711 | 11726 | . 2683 |  | . 2726 |
| 1991 | 2717 | 38662 | 34663 | 8429 | . 2432 |  | . 2752 |
| 1992 | 9452 | 28997 | 26892 | 5476 | . 2036 |  | . 2109 |
| 1993 | 132839 | 25784 | 23067 | 4026 | . 1745 |  | . 1889 |
| 1994 | 59389 | 27222 | 21427 | 4252 | . 1984 |  | . 2073 |
| 1995 | 11228 | 82884 | 22389 | 4948 | . 2210 |  | . 2330 |
| 1996 | 4473 | 101356 | 47520 | 9642 | . 2029 |  | . 3269 |
| 1997 | 22697 | 97423 | 75143 | 17924 | . 2385 |  | . 3860 |
| 1998 | 34346 | 82007 | 73269 | 22210 | . 3031 |  | . 5665 |
| 1999 | 145443 | 62829 | 53552 | 18482 | . 3451 |  | . 5561 |
| 2000 | 70924 | 64457 | 42004 | 15821 | . 3767 |  | . 3671 |
| 2001 | 70651 | 111963 | 49741 | 16339 | . 3285 |  | . 4015 |
| 2002 | 36965 | 132160 | 73029 | 25584 | . 3503 |  | . 4499 |
| Arith. |  |  |  |  |  |  |  |
| Mean | 42395 | 81150 | 56997 | 16219 | . 2907 |  | .3508 |
| 0 Units | (Thousands) | (Tonnes) | (Tonnes) | (Tonnes) |  |  |  |

See Table 2.4.18 for the summary table with recruitment at age 2 !!

Stock size
The yearclasses up to 2000 included are derived from the final 2003 VPA
The yearclasses 2001-2002 at age 2 are estimated using XSA tuned with surveys including indices for age 0-2
and apply a natural mortality of 0.2 in formard calculations of the numbers using standard VPA equations
The yearclass 2003 at age 2 in 2005 is estimated as the geomean of the yearclasses since 1980


Catch/stock weights at age


Exploitation pattern

| Age | 2003 | 2004 | 2005 | 20100 | 20101 | 2002 | Awerage $F$ for 2000-02 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.0540 | 0.0540 | 0.0540 | 0.0727 | 0.0524 | 0.0368 | 2 | 0.0540 |
| 3 | 0.2614 | 0.2614 | 0.2614 | 0.3217 | 0.2266 | 0.236 | 3 | 0.2614 |
| 4 | 0.3695 | 0.3695 | 0.3695 | 0.299 | 0.4734 | 0.336 | 4 | 0.3695 |
| 5 | 0.4234 | 0.4234 | 0.4234 | 0.3677 | 0.4476 | 0.4548 | 5 | 0.4234 |
| 6 | 0.4689 | 0.4689 | 0.4689 | 0.4769 | 0.4609 | 0.7485 | 6 | 0.4689 |
| 7 | 0.4146 | 0.4146 | 0.4146 | 0.3704 | 0.3992 | 0.4741 | 7 | 0.4146 |
| 8 | 0.4467 | 0.4467 | 0.4467 | 0.7825 | 0.3175 | 0.5758 | 8 | 0.4467 |
| 9 | 0.4206 | 0.4206 | 0.4206 | 0.4373 | 0.3739 | 0.4506 | 9 | 0.4206 |
| $10+$ | 0.4206 | 0.4206 | 0.4206 | 0.4373 | 0.3739 | 0.4506 | $10+$ | 0.4206 |
| Avg3-7 | 0.3875 | 0.3875 | 0.3875 | 0.3671 | 0.4015 | 0.4499 | Fbar(3-7) | 0.3875 |

The exploitation pattern is estimated from the average fishing mortality matrix $2000-2002$ from the final VPA in 2002.
The high $F$ values for age 6 in 2000 and for age 6 in 2002 have been excluded from the aweraging

Table 2.4.14

MFDP version 1
Run: farhad-mopt1
Time and date: 14:56 5/7/03
Fbar age range: 3-7

| 2003 <br> Age |  | N | M | Mat | PF | PM | SWt | Sel |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | CWt | CW |
| ---: | ---: |


| 2004 <br> Age | $\mathbf{N}$ | $\mathbf{M}$ | Mat | PF | PM | SWt | Sel | CWt |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 24800 | 0.2 | 0.1 | 0 | 0 | 0.584 | 0.0540 | 0.584 |
| 3. |  | 0.2 | 0.5 | 0 | 0 | 0.805 | 0.2614 | 0.805 |
| 4. |  | 0.2 | 0.96 | 0 | 0 | 1.050 | 0.3695 | 1.050 |
| 5. | 0.2 | 0.99 | 0 | 0 | 1.452 | 0.4234 | 1.452 |  |
| 6. | 0.2 | 1 | 0 | 0 | 2.198 | 0.4689 | 2.198 |  |
| 7. | 0.2 | 1 | 0 | 0 | 2.358 | 0.4146 | 2.358 |  |
| 8. | 0.2 | 1 | 0 | 0 | 2.288 | 0.4467 | 2.288 |  |
| 9. | 0.2 | 1 | 0 | 0 | 2.520 | 0.4206 | 2.520 |  |
| 10. | 0.2 | 1 | 0 | 0 | 3.130 | 0.4206 | 3.130 |  |


| 2005 <br> Age | $\mathbf{N}$ | $\mathbf{M}$ | Mat | PF | PM | SWt | Sel | CWt |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 14300 | 0.2 | 0.1 | 0 | 0 | 0.584 | 0.0540 | 0.584 |
| 3. |  | 0.2 | 0.5 | 0 | 0 | 0.805 | 0.2614 | 0.805 |
| 4. |  | 0.2 | 0.96 | 0 | 0 | 1.050 | 0.3695 | 1.050 |
| 5. | 0.2 | 0.99 | 0 | 0 | 1.364 | 0.4234 | 1.364 |  |
| 6. | 0.2 | 1 | 0 | 0 | 1.748 | 0.4689 | 1.748 |  |
| 7. | 0.2 | 1 | 0 | 0 | 2.393 | 0.4146 | 2.393 |  |
| 8. |  | 0.2 | 1 | 0 | 0 | 2.510 | 0.4467 | 2.510 |
| 9. | 0.2 | 1 | 0 | 0 | 2.354 | 0.4206 | 2.354 |  |
| 10. | 0.2 | 1 | 0 | 0 | 3.130 | 0.4206 | 3.130 |  |

Input units are thousands and kg - output in tonnes

Table 2.4.15 Faroe Haddock. Management option table - Results
MFDP version 1
Run: farhad-mopt1
Index file 05/05/2003
Time and date: 14:56 5/7/03
Fbar age range: 3-7
2003

Biomass $\quad$ SSB $\quad$ FMult | FBar Landings |  |  |
| ---: | :--- | ---: |
| 138400 | 96294 | 1 | $0.3875 \quad 31294$

| 2004 |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Biomass | SSB | FMult | FBar | Landings | 2005 |  |
| 124790 | 95560 | 0 | 0 | 0 | 135975 | 118758 |
| . | 95560 | 0.1 | 0.0388 | 3709 | 131877 | 114746 |
| . | 95560 | 0.2 | 0.0775 | 7286 | 127930 | 110884 |
| . | 95560 | 0.3 | 0.1163 | 10734 | 124128 | 107165 |
| . | 95560 | 0.4 | 0.155 | 14060 | 120464 | 103583 |
| . | 95560 | 0.5 | 0.1938 | 17267 | 116934 | 100134 |
| . | 95560 | 0.6 | 0.2325 | 20360 | 113532 | 96812 |
| . | 95560 | 0.7 | 0.2713 | 23344 | 110255 | 93612 |
| . | 95560 | 0.8 | 0.31 | 26223 | 107096 | 90530 |
| . | 95560 | 0.9 | 0.3488 | 29000 | 104051 | 87561 |
| . | 95560 | 1 | 0.3875 | 31679 | 101116 | 84701 |
| . | 95560 | 1.1 | 0.4263 | 34265 | 98288 | 81946 |
| . | 95560 | 1.2 | 0.4651 | 36760 | 95561 | 79291 |
| . | 95560 | 1.3 | 0.5038 | 39168 | 92931 | 76733 |
| . | 95560 | 1.4 | 0.5426 | 41492 | 90397 | 74268 |
| . | 95560 | 1.5 | 0.5813 | 43736 | 87952 | 71893 |
| . | 95560 | 1.6 | 0.6201 | 45902 | 85596 | 69605 |
| . | 95560 | 1.7 | 0.6588 | 47993 | 83323 | 67399 |
| . | 95560 | 1.8 | 0.6976 | 50012 | 81130 | 65273 |
| . | 95560 | 1.9 | 0.7363 | 51963 | 79016 | 63223 |
| . | 95560 | 2 | 0.7751 | 53846 | 76976 | 61248 |

Input units are thousands and kg - output in tonnes
Table 2.4.16 Faroe Haddock Long-term Prediction Input data
MFYPR version 1
Run: farhad-ypr1
Index file 05/05/2003
Time and date: 15:22 5/7/03
Fbar age range: 2-10

| Age | M | Mat | PF | PM | SWt | Sel | CWt |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.2 | 0.06 | 0 | 0 | 0.558 | 0.0759 | 0.558 |
| 3 | 0.2 | 0.46 | 0 | 0 | 0.823 | 0.2650 | 0.823 |
| 4 | 0.2 | 0.91 | 0 | 0 | 1.106 | 0.3695 | 1.106 |
| 5 | 0.2 | 0.99 | 0 | 0 | 1.443 | 0.3811 | 1.443 |
| 6 | 0.2 | 1.00 | 0 | 0 | 1.757 | 0.4558 | 1.757 |
| 7 | 0.2 | 1.00 | 0 | 0 | 2.052 | 0.5776 | 2.052 |
| 8 | 0.2 | 1.00 | 0 | 0 | 2.260 | 0.6538 | 2.260 |
| 9 | 0.2 | 1.00 | 0 | 0 | 2.504 | 0.4923 | 2.504 |
| 10 | 0.2 | 1.00 | 0 | 0 | 2.836 | 0.4923 | 2.836 |

Table 2.4.17 Faroe Haddock Long -term Prediction Results
MFYPR version 1
Run: farhad-ypr1
Time and date: 15:22 5/7/03
Yield per results

| FMult |  | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | SpwnNosSpwn | SSBSpwn |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 0.0000 | 0.0000 | 0.0000 | 5.5167 | 8.7665 | 4.0696 | 7.8024 | 4.0696 | 7.8024 |  |
|  | 0.1 | 0.0418 | 0.1480 | 0.2633 | 4.7798 | 6.9250 | 3.3386 | 5.9667 | 3.3386 | 5.9667 |
|  | 0.2 | 0.0836 | 0.2475 | 0.4122 | 4.2849 | 5.7427 | 2.8496 | 4.7900 | 2.8496 | 4.7900 |
|  | 0.3 | 0.1254 | 0.3194 | 0.5012 | 3.9281 | 4.9284 | 2.4983 | 3.9812 | 2.4983 | 3.9812 |
|  | 0.4 | 0.1673 | 0.3741 | 0.5565 | 3.6571 | 4.3378 | 2.2328 | 3.3959 | 2.2328 | 3.3959 |
|  | 0.5 | 0.2091 | 0.4174 | 0.5917 | 3.4431 | 3.8919 | 2.0243 | 2.9552 | 2.0243 | 2.9552 |
|  | 0.6 | 0.2509 | 0.4527 | 0.6144 | 3.2690 | 3.5443 | 1.8555 | 2.6126 | 1.8555 | 2.6126 |
|  | 0.7 | 0.2927 | 0.4822 | 0.6292 | 3.1239 | 3.2661 | 1.7155 | 2.3393 | 1.7155 | 2.3393 |
|  | 0.8 | 0.3345 | 0.5073 | 0.6387 | 3.0006 | 3.0384 | 1.5973 | 2.1164 | 1.5973 | 2.1164 |
|  | 0.9 | 0.3763 | 0.5290 | 0.6447 | 2.8941 | 2.8485 | 1.4958 | 1.9313 | 1.4958 | 1.9313 |
| 1 | 0.4181 | 0.5480 | 0.6483 | 2.8010 | 2.6878 | 1.4075 | 1.7752 | 1.4075 | 1.7752 |  |
|  | 1.1 | 0.4599 | 0.5649 | 0.6503 | 2.7185 | 2.5498 | 1.3299 | 1.6417 | 1.3299 | 1.6417 |
|  | 1.2 | 0.5018 | 0.5800 | 0.6511 | 2.6449 | 2.4300 | 1.2609 | 1.5263 | 1.2609 | 1.5263 |
|  | 1.3 | 0.5436 | 0.5936 | 0.6510 | 2.5786 | 2.3249 | 1.1993 | 1.4255 | 1.1993 | 1.4255 |
|  | 1.4 | 0.5854 | 0.6059 | 0.6504 | 2.5186 | 2.2319 | 1.1437 | 1.3368 | 1.1437 | 1.3368 |
| 1.5 | 0.6272 | 0.6172 | 0.6495 | 2.4638 | 2.1490 | 1.0934 | 1.2580 | 1.0934 | 1.2580 |  |
|  | 1.6 | 0.6690 | 0.6276 | 0.6482 | 2.4135 | 2.0745 | 1.0476 | 1.1876 | 1.0476 | 1.1876 |
| 1.7 | 0.7108 | 0.6372 | 0.6467 | 2.3673 | 2.0073 | 1.0056 | 1.1243 | 1.0056 | 1.1243 |  |
| 1.8 | 0.7526 | 0.6461 | 0.6451 | 2.3245 | 1.9462 | 0.9670 | 1.0671 | 0.9670 | 1.0671 |  |
|  | 1.9 | 0.7945 | 0.6543 | 0.6434 | 2.2848 | 1.8905 | 0.9314 | 1.0152 | 0.9314 | 1.0152 |
| 2 | 0.8363 | 0.6620 | 0.6417 | 2.2478 | 1.8395 | 0.8985 | 0.9679 | 0.8985 | 0.9679 |  |


| Reference point | F multiplier Absolute F |  |
| :--- | ---: | ---: |
| Fbar(2-10) | 1.0000 | 0.4181 |
| FMax | 1.2461 | 0.5211 |
| F0.1 | 0.4472 | 0.1870 |
| F35\%SPR | 0.5628 | 0.2353 |
| Flow | -99 |  |
| Fmed | 0.5520 | 0.2308 |
| Fhigh | 1.9955 | 0.8344 |

Weights in kilograms

Table 2.4.18 Faroe haddock (Division Vb) Stock Summary Table

| Year | Recruitment <br> Age 2 <br> thousands | SSB <br> tonnes | Landings <br> tonnes | Mean F <br> Ages 3-7 |
| :---: | :---: | :---: | :---: | :---: |
| 1961 | 51279 | 47797 | 20831 | 0.5624 |
| 1962 | 38537 | 51875 | 27151 | 0.6506 |
| 1963 | 47362 | 49547 | 27571 | 0.7002 |
| 1964 | 30110 | 44128 | 19490 | 0.4753 |
| 1965 | 22644 | 45555 | 18479 | 0.5260 |
| 1966 | 20206 | 43953 | 18766 | 0.5288 |
| 1967 | 25356 | 41959 | 13381 | 0.4030 |
| 1968 | 54843 | 45379 | 17852 | 0.4377 |
| 1969 | 31968 | 53422 | 23272 | 0.4853 |
| 1970 | 35582 | 59858 | 21361 | 0.4762 |
| 1971 | 15450 | 62908 | 19393 | 0.4564 |
| 1972 | 33176 | 61975 | 16485 | 0.3964 |
| 1973 | 23690 | 61578 | 17976 | 0.2894 |
| 1974 | 52334 | 64631 | 14773 | 0.2206 |
| 1975 | 70053 | 75405 | 20715 | 0.1799 |
| 1976 | 55969 | 89220 | 26211 | 0.2475 |
| 1977 | 26191 | 96373 | 25555 | 0.3873 |
| 1978 | 35099 | 97226 | 19200 | 0.2781 |
| 1979 | 2784 | 85394 | 12418 | 0.1551 |
| 1980 | 4944 | 81902 | 15016 | 0.1779 |
| 1981 | 3491 | 75846 | 12233 | 0.1813 |
| 1982 | 15835 | 56804 | 11937 | 0.3308 |
| 1983 | 19615 | 51811 | 12894 | 0.2654 |
| 1984 | 40781 | 53820 | 12378 | 0.2284 |
| 1985 | 39422 | 62602 | 15143 | 0.2761 |
| 1986 | 26448 | 65604 | 14477 | 0.2237 |
| 1987 | 9421 | 67294 | 14882 | 0.2643 |
| 1988 | 18733 | 61882 | 12178 | 0.2010 |
| 1989 | 14464 | 51703 | 14325 | 0.2854 |
| 1990 | 9048 | 43711 | 11726 | 0.2726 |
| 1991 | 2927 | 34663 | 8429 | 0.2752 |
| 1992 | 2660 | 26892 | 5476 | 0.2109 |
| 1993 | 1821 | 23067 | 4026 | 0.1889 |
| 1994 | 6336 | 21427 | 4252 | 0.2073 |
| 1995 | 89045 | 22389 | 4948 | 0.2330 |
| 1996 | 39810 | 47520 | 9642 | 0.3269 |
| 1997 | 7526 | 75143 | 17924 | 0.3860 |
| 1998 | 2998 | 73269 | 22210 | 0.5665 |
| 1999 | 15214 | 53552 | 18482 | 0.5561 |
| 2000 | 23023 | 42004 | 15821 | 0.3671 |
| 2001 | 97493 | 49741 | 16339 | 0.4015 |
| 2002 | 47542 | 73029 | 25584 | 0.4499 |
| 2003 | 47400 | 96290 |  | 0.3875 |
| Average | 29270 | 57910 | 16219 | 0.3516 |

Yield and spawning biomass per Recruit
F-reference points:

|  | Fish Mort <br> Ages 3-7 | Yield/R | SSB/R |
| :--- | :---: | :---: | :---: |
| Average Current | 0.406 | 0.648 | 1.788 |
| Fmax | 0.511 | 0.651 | 1.478 |
| F0.1 | 0.183 | 0.575 | 3.173 |



Figure 2.4.1 Faroe Haddock. Landings by all nations since 1903.


Figure 2.4.2 Faroe Haddock. Cumulative landings



Figure 2.4.3
Faroe Haddock. Faroese landings from A) Vb1 and B) Vb2 in 2002 by fleet. Tonnes ungutted weight.


Figure 2.4.4 Faroe Haddock. LN (catch-at-age in numbers) for YC's 1953 onwards.

## Faroe haddock weight at age



Figure 2.4.5 Faroe Haddock. Mean weight-at-age (2-7).


Figure 2.4.6
Faroe Haddock. Mean weight-at-age Jan-Feb.


A: Faroe haddock. Maturity ogives. Observed values from the spring survey.


B: Faroe haddock. Maturity ogives. Running 3 years average from the spring survey.

Figure 2.4.7 Faroe Haddock. Maturity-at-age.


Figure 2.4.8A. Faroe Haddock.


Figure 2.4.8B. Faroe Haddock.


Figure 2.4.9A. Faroe Haddock.


Figure 2.4.9B. Faroe Haddock.

Faroe Haddock Summer Survey


Figure 2.4.10 Faroe haddock. LN(catch-at-age in numbers).



Figure 2.4.11 Log q Residuals from the accepted runs. Upper figure is the summer survey and the lower figure is the spring survey.


Figure 2.4.12 Relationship between indices age a and age $\mathrm{a}+1$ for the same year class in the summer survey


Figure 2.4.13 Faroe haddock. LN(catch-at-age in numbers).









Figure 2.4.14 Relationship between indices age a and age $\mathrm{a}+1$ for the same year class in the spring survey

## Faroe haddock tuning series



Figure 2.4.16 Faroe haddock. Age-aggregated tuning series.


Figure 2.4.17 Faroe haddock. Retrospective analysis


Figure 2.4.18 Comparison of the four different XSA runs shown in Figure 2.4.17.


Figure 2.4.19 Faroe haddock. Comparison of 11 different XSA runs.


MFYPR version 1
Run: farhad-ypr1
Time and date: 15:22 5/7/03

|  |  |  |
| :--- | :---: | :---: |
| Reference point | F multiplier | Absolute F |
| Fbar(2-10) | 1.0000 | 0.4181 |
| FMax | 1.2461 | 0.5211 |
| F0.1 | 0.4472 | 0.1870 |
| F53\%SPR | 0.5628 | 0.2353 |
| Flow | -99.0000 |  |
| Fmed | 0.5520 | 0.2308 |
| Fhigh | 1.9955 | 0.8344 |

Weights in kilograms


MFDP version 1
Run: farhad-mopt1
Index file 05/05/2003
Time and date: 14:56 5/7/03
Fbar age range: 3-7
Input units are thousands and kg - output in tonnes

Figure 2.4.21. Faroe haddock predictions


Figure 2.4.22
Faroe haddock stock recruitment plot.




Figure 2.4.23
Faroe haddock (Division Vb ). In the text the figures are labelled A-G




Figure 2.4.23 (Cont'd)

### 2.5.1 Landings and trends in the fishery

Nominal landings of saithe from the Faroese grounds (Division Vb) have been highly variable since 1960 ranging from 10000 t to 60000 t over that period. In 1990 record high landings of about 60000 t were taken. Thereafter landings declined steadily to 20000 t in 1996. Since then landings have increased to 39000 t in 2000, 51800 tonnes in 2001 and to 56700 tonnes in 2002 (Table 2.5.1.1, Figure 2.5.1.1).

With the introduction of the 200 miles EEZ in 1977, saithe has mainly been fished by Faroese vessels. The principal fleet consists of large pair trawlers ( $>1000 \mathrm{HP}$ ), which have a directed fishery for saithe, accounting for about $60 \%$ of the reported landings in 1993-2002 (Table 2.5.1.2). The smaller pair trawlers ( $<1000 \mathrm{HP}$ ) have a more mixed fishery and they account for about $10-20 \%$ of the total landings of saithe in 1993-2002. During the last decade the proportion of saithe in the catches has generally increased for larger pair trawlers and larger single trawlers ( $>1000 \mathrm{HP}$ ) but decreased for the smaller pair trawlers and jiggers. Other vessels only have small catches of saithe as bycatch.

Catches used in the assessment are presented in Table 2.5.1.1. These include foreign catches that have been reported to the Faroese Authorities but not officially reported to ICES. Catches in that part of Subdivision IIa which lies immediately north of the Faroes have also been included. Little discarding is thought to occur in this fishery.

### 2.5.2 Catch-at-age

Catch-at-age is based on length and otolith samples from Faroese landings of jiggers, small and large single and pair trawlers, and landing statistics by fleet provided by the Faroese Authorities. Catch-at-age was calculated for each fleet by four-month periods, before the numbers were combined. Catch-at-age was thereafter raised by the foreign catches. The catch-at-age data for previous years were also revised according to the final catch statistics (Tables 2.5.2.1 and 2.5.2.2). The sampling intensity in 2002 was similar to that in 2001:

| Fleet | Samples | Lengths | Otoliths | Weights |
| :--- | ---: | ---: | ---: | ---: |
| Jiggers | 5 | 1197 | 120 | 120 |
| Single trawlers $1000-1499 \mathrm{HP}$ | 3 | 641 | 61 | 0 |
| Single trawlers $1500-1999 \mathrm{HP}$ | 16 | 4037 | 480 | 360 |
| Single trawlers > 2000 HP | 18 | 4557 | 750 | 60 |
| Pair trawlers $400-699 \mathrm{HP}$ | 4 | 975 | 120 | 120 |
| Pair trawlers 700 - 999 HP | 16 | 4074 | 302 | 120 |
| Pair trawlers 1000 - 1499 HP | 127 | 30761 | 3001 | 2760 |
| Total | $\mathbf{1 8 9}$ | $\mathbf{4 6 2 4 2}$ | $\mathbf{4 8 3 4}$ | $\mathbf{3 5 4 0}$ |

### 2.5.3 Weight-at-age

Mean weights-at-age have varied by a factor of about 2 during 1961-2002. For example, the mean weights-at-age 5 varied between about 1.6 kg in 1973 and 3.3 kg in 1980, while at age 7 it varied between 2.6 kg in 1991 and 5.3 kg in 1985 (Table 2.5.3.1 and Figure 2.5.3.1). Mean weights-at-age were generally high during 1980-86 and dropped in the period 1987-1991. The mean weights increased again in the period 1992-96 but have shown a general decrease since then. The SOP for 2002 was $100 \%$.

### 2.5.4 Maturity-at-age

Maturity-at-age data is available from 1983 onward. Due to poor sampling in 1988 the proportion mature for that year was calculated as the average of the two adjacent years. A model was used, described in the 1993 Working Group report (ICES C.M.1993/Assess:18), for predicting maturity-at-age in order to alleviate some of the problems involved with the sampling data. The basic model used was a GLM with a Logit link function describing maturity-at-age as a function of age, year-class strength, mean weight-at-age and a year effect. This model was applied to predict the entire maturity-at-age for 1983-2002 (Table 2.5.4.1 and Figure 2.5.4.1).

For the GLM maturity model, weight is the mean weight-at-age in the catch and year-class strength is from the final XSA run in assessment year 2002. Because ages 3 and 4 are generally not derived directly from the assessment, the following estimates are used in XSA runs:

- Year-class strength at age 3 is the geometric mean for period 1983-2002
- Year-class strength at age 4 is the recruitment estimate of age 3 in year 2002.

In 2003 the maturity data for the period 1994-1997 were corrected according to WD 14, NWWG 2003. The GLM model was used on these corrected data. The maturity estimates from the model based on the corrected data are essentially identical to those using the uncorrected data.

### 2.5.5 Stock assessment

### 2.5.5.1 Tuning and estimation of fishing mortality

The summer survey (1996-2002), similar to last year, showed large standard errors of $\log \mathrm{q}$, and there was a marked trend in residuals. The results of a spring survey (1994-2002) were investigated but showed poor internal consistency. The Working Group re-iterates its recommendation that the survey data be further investigated to improve the usefulness of the survey series as indices of stock size.

The single CPUE series used in the assessment since 2000 was introduced in 1998 (ICES C.M. 1998/ACFM:19), and consists of saithe catch-at-age and effort in hours, referred to as the Cuba Logbook series. The series extends back to 1985 and consists of data from 8 pair trawlers greater than 1000 HP (Cuba trawlers) which specialize in fishing on saithe and account for $5000-10000 \mathrm{t}$ of saithe each year (described in annex). In the Cuba Logbook series, information for each haul was supplied and only those hauls where saithe contributed more than $50 \%$ of the total catches of cod, haddock and saithe were used (Table 2.5.5.1).

The Working Group was concerned that the catchability of the tuning series may have changed over time. Catchability was estimated by dividing the CPUE (in kg per hour) by exploitable biomass, that is biomass multiplied by a selection pattern reaching a value of 1 for the fully exploited age groups. Two sets of smoothed selection pattern were used: one estimated by a five-year period, the other estimated for two periods: 1961 to 1979 and 1980 to 2002. The exploitation pattern shows an increasing trend from 1991 to 1996, but the estimates have been reasonably stable for the period 19972002 (shown in the figures below). The estimates, however, are calculated from an assessment calibrated with the CPUE series, and recent values are dependent on the CPUE series. The Working Group accepted the XSA calibrated with the Cuba trawler CPUE series, considering that no other usable index of stock size is aviable.


The final XSA run in the current assessment was made with the same parameters as in 2002. The CPUE series used are shown in Table 2.5.5.1. The XSA diagnostics are in Table 2.5.5.2 and the output from the XSA is presented in Tables 2.5.5.3-5. The values of S.E. $(\log q)$ are high, but for the principal year classes they appear reasonable. The logcatchability residuals from the XSA tuning for age groups 3 and 5-11 (Figure 2.5.5.1) show more negative values in the first eight years than in the last nine years of the 17 -year time-series.

Retrospective analysis of the average fishing mortality for age groups $4-8$ years (Figure 2.5.5.2) shows a tendency to overestimate F . This implies that biomass was correspondingly underestimated (Figure 2.5.5.3).

The fishing mortalities for 1961-2002 are presented in Table 2.5.5.3 and in Figure 2.5.5.4. The average fishing mortality for age groups 4-8 was 0.38 in 2002.

### 2.5.5.2 Stock estimates and recruitment

Recruitment in the 1980s was above or close to average ( 28 millions). The strongest year class since 1986 was produced in the 1990s and the average for that decade is about 39 million (Figure 2.5.5.5). The 1998 year class is unusually large as can be seen in the modal length progression in the summer survey from 1999 to 2002 (Figure 2.5.5.8). Even though recruitment had been above average in the 1960s and 1970s, SSB declined from nearly 112000 t in 1985 to 71000 t in 1991 as a result of high fishing mortality yielding the highest (1990) and third highest (1996) landings of the whole 1961-2001 period. The historically low SSB persisted in 1992-1995 (Table 2.5.5.5 and Figure 2.5.5.6). The SSB has increased since 1996 with the maturation of the 1992, 1994, 1996 and 1998 year classes. SSB was estimated to be 122000 t in 2002, which is about the same as 2001, and well above the average SSB ( 94000 t ) in the 1980s. The relation between stock and recruitment is shown in Figure 2.5.5.7.

### 2.5.6 Prediction of catch and biomass

### 2.5.6.1 Input data

Input data for prediction with management options are presented in Table 2.5.6.1 and input data for the yield-per-recruit calculations are given in Table 2.5.6.2.

The size of the 1998 year class from the calibration was higher than the highest observed. Both the survey and the commercial cpues indicate that this year class is very abundant (Figure 2.5.5.8). The size was reset to approximately the highest observed ( 1996 year class at age 3 in $1999=80000$ ) and the fishing mortality in 2001 on those year classes was adjusted accordingly. There was also a projection with management options done based on the predicted 1998 year class from the XSA assessment.

| Year | YC | XSA | Value used | Basis |
| :---: | :---: | :---: | :---: | :---: |
| $\mathbf{2 0 0 1}$ | $\mathbf{1 9 9 8}$ | 103872 | 80000 | Max Obs |
| $\mathbf{2 0 0 2}$ | $\mathbf{1 9 9 9}$ | 29540 | 29540 | XSA |
| $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 0}$ |  | 29650 | Geomean77-99yc |
| $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 1}$ |  | 29650 | Geomean77-99yc |
| $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 2}$ |  | 29650 | Geomean77-99yc |

Population numbers for the short-term prediction up to the 1999 year class are from the final VPA run, whereas values for the 2000-2002 year classes are the geometric mean of the 1977 to 1999 year classes. Mean weights for the stock and for the catches are the same for 2003-2005, the arithmetic mean for 2000-2002. In the long-term prediction (yield-perrecruit) mean weights for 1961-2002 were used.

In the short-term prediction the fitted proportion mature values from the model for 2003 were used for that year and for 2004 and 2005 the average of fitted values for 2001-2003 was used. In the long-term prediction the average of fitted values for 1983-2003 was used.

For all three years in the short-term prediction the average exploitation pattern in the final VPA for 2000-2002, not rescaled to Fbar (ages 4-8) in 2002 in view of a retrospective problem (as suggested from ACFM), was used. In the long-term prediction the exploitation pattern was set equal to the average of exploitation patterns for 1961-2002.

### 2.5.6.2 Biological reference points

Yield-per-recruit and spawning stock biomass per recruit curves are presented in Figure 2.5.6.1. Compared to the 2002 average fishing mortality of 0.33 in age groups $4-8, \mathbf{F}_{\max }$ is $0.42, \mathbf{F}_{0.1}$ is $0.16, \mathbf{F}_{\text {med }}$ is 0.35 and $\mathbf{F}_{\text {high }}$ is 0.82 (Table 2.5.6.3, Figure 2.5.6.1 and Figure 2.5.6.2).

Yield and spawning biomass per Recruit F-reference points:

|  | Fish Mort <br> Ages 4-8 | Yield/R | SSB/R |
| :--- | :---: | :---: | :---: |
| Average last 3 years | 0.413 | 1.494 | 3.112 |
| $\mathbf{F}_{\text {max }}$ | 0.418 | 1.494 | 3.072 |
| $\mathbf{F}_{0.1}$ | 0.158 | 1.330 | 7.073 |
| $\mathbf{F}_{\text {med }}$ | 0.336 | 1.488 | 3.787 |

ACFM set $\mathbf{F}_{\text {lim }}=0.40$ and $\mathbf{F}_{\mathrm{pa}}=0.28$ (May 1998), and $\mathbf{B}_{\mathrm{lim}}=60000 \mathrm{t}$ and $\mathbf{B}_{\mathrm{pa}}=85000 \mathrm{t}$ (May 1999). The current assessment (Table 2.5.5.5 and Figure 2.5.5.4) shows that fishing mortality has averaged 0.33 over the time period, that $F$ has been above $\mathbf{F}_{\text {lim }}$ in 13 out of 42 years ( $30 \%$ ). Fishing mortality of $F=0.40$ therefore does not appear to be associated with a high probability of stock collapse. Fishing mortality has been above $\mathbf{F}_{\mathrm{pa}}$ every year except one since the 1980. There does appear, however, to be some relationship between fishing mortality and SSB with lower SSBs associated with higher fishing mortality (Figure 2.5.6.3). When fishing mortality and SSB are plotted versus time (Figure 2.5.5.4 and 2.5.5.6), initially, both F and SSB increase at the same time. SSB peaks in 1972 and subsequently decreases until 1981 while fishing mortality continues to increase. The brief increase in SSB in the early 1980s occurred at F in the order of 0.40 . The 1983 to 1985 year classes were all relatively strong. Yet they did not result in substantial increases in SSB, because of high fishing mortality. Medium-term simulations done during the 2001 NWWG meeting showed either stable SSB or slightly decreasing SSBs at F status quo (0.41) with a negligible probability that SSB would fall below $\mathbf{B}_{\text {lim }}$. Therefore, fishing mortalities in the order of 0.33 , the average over the available time-series, do not appear associated with a high probability of stock collapse as implied by $\mathbf{F}_{\text {lim }}$. Given the history of the stock and the possible influence of changes in productivity on Faroese stocks, $\mathrm{F}=0.33$ could be considered as $\mathbf{F}_{\mathrm{pa}}$, with $\mathbf{F}_{\text {lim }}$ derived using the usual formula of $\mathbf{F}_{\text {lim }}=\mathrm{F}_{\mathrm{pa}} \mathrm{e}^{1.645 \mathrm{xs}}$ where s could be 0.40 resulting in an $\mathbf{F}_{\text {lim }}=0.64$, a value exceeded only once in the history of the fishery. The SG on Precautionary Reference Points for Advice on Fishery Management (SGPRP February 2003) suggested that the existing $\mathbf{B}_{\text {lim }}$ for Faroe saithe could be a candidate for $\mathbf{B}_{\mathrm{pa}}$ instead. Considering that ACFM is unlikely to define and use new $\mathbf{B}_{\text {lim }}$ points, the Working Group will consider the issue further during its 2004 meeting.

The stock and recruitment relationship for Faroe saithe (Figure 2.5.5.7) shows a pattern somewhere in-between types 2a and 3a identified by the SGPA, i.e. generally increasing recruitment as SSB decreases with some smaller recruitment at the lowest SSB. Given this pattern, it is not possible to identify with any certainty the SSB below which recruitment becomes impaired. Using the lowest SSB as $\mathbf{B}_{\text {lim }}$ would probably be overly conservative, while using it as $\mathbf{B}_{\mathrm{pa}}$ may be somewhat risky. Given the difficulties in identifying the biomass where recruitment becomes impaired, the equilibrium SSB corresponding to $\mathbf{F}_{\mathrm{pa}}$ above could be used as $\mathbf{B}_{\mathrm{pa}}$ and $\mathbf{B}_{\mathrm{lim}}$ could be calculated using the usual formula $\mathbf{B}_{\mathrm{lim}}=\mathrm{B}_{\mathrm{pa}} \mathrm{e}^{1.645 \mathrm{xs}}$ where s could be 0.30 .

The history of the stock/fishery in relation to the four reference points can be seen in Figure 2.5.6.3.

### 2.5.6.3 Projection of catch and biomass

Results from predictions with management option are presented in Table 2.5.6.4 and Figure 2.5.6.1. Catches at status quo F would be 67000 t in 2003 and 66000 t in 2004. The spawning stock biomass would be about 1.7 times higher than $\mathbf{B}_{\mathrm{pa}}$ in 2002 and 2003.

Results from the yield-per-recruit estimates are shown in Table 2.5.6.3 and Figure 2.5.6.1.

A projection of catch in number by year classes in 2003 and weight composition in SSB by year classes in 2004 is presented in Figure 2.5.6.4. The catch in 2003 is predicted to rely on the four most recent year classes ( $84 \%$ ). In 2004 the 1996, 1997 and 1998 year classes are expected to contribute over $80 \%$ of the SSB.

### 2.5.7 Management considerations

The spawning stock biomass has increased to above $\mathbf{B}_{\mathrm{pa}}$ and is expected to remain above $\mathbf{B}_{\mathrm{pa}}$ at status quo fishing mortality, due to good recruitment in the short term.

### 2.5.8 Comments on the assessment

The XSA settings and tuning fleets are the same as last year.
There is still no independent recruitment index to predict recruits in the first year in the short-term prediction. Attempts have been made to establish a programme for echo sounding and biological sampling of age group 0-2. However, this needs to be developed further and consequently no results are available at this stage. It has been suggested by NWWG that an attempt should be tried to analyse the correlation between survey index and stock in number from VPA, principally ages 2 and 3 .

The question of migration has been brought up previously. Although tagging data indicates that saithe migrate between management areas, and some indications are seen in the assessment as well, no attempts have been made to quantify the migration rate of saithe. An analysis of saithe otoliths using otolith elemental fingerprinting (OEF) between
management areas in the North Atlantic is initiated by Iceland, which will hopefully add valuable information on saithe stocks in the north Atlantic and migrations between management units.

The 2003 assessment has been calibrated in a way very similar to the 2002 assessment. The results indicate that the point estimator of biomass is lower than was estimated in the 2002 assessment ( $2002 \mathrm{SSB}=120000 \mathrm{t}$ compared with 140000 t ) and that fishing mortality is almost the same. In the 2001 assessment, recent year classes were assumed to be equal to a gm mean of recent year classes. In the 2003 assessment, the estimates from the XSA calibration, adjusted down to the highest previously observed, were used.

### 2.5.9 Annex

## Stock definition

Saithe are widely distributed around the Faroes, from the shallow inshore waters to depths of 350 m . The main spawning areas are found at 150-250 meters depth east and north of the Faroes. Spawning takes place from January to April, with the main spawning in the second-half of February. The pelagic eggs and larvae drift with the anti-cyclonic current around the islands until May/June, when the juveniles, at lengths of $2.5-3.5 \mathrm{~cm}$, migrate inshore. The nursery areas during the first two years of life are in very shallow waters in the littoral zone. Young saithe are also distributed in shallow depths, but at increasing depths with increasing age. Saithe enter the adult stock at the age of 3 or 4 years (Jákupsstovu 1999). Tagging experiments of saithe has demonstrated migrations between the Faroes, Iceland, Norway, west of Scotland and the North Sea (Jákupsstovu 1999).

## Description of the Cuba pair-trawlers

The tuning fleet called Cuba-trawlers consists of trawlers that were built in East-Germany in 1970 as part of a helpprogramme for Cuba (explaining the name). In 1973 "Faroe Ship" bought 8 of these trawlers and brought them to Faroe Islands. Today they are kept by the Runavik Trawl Company "Beta", which is the company that has operated the trawlers during all these years and has registered the catches.

The Cuba-fleet first operated in the North Sea as a standby supply service for drilling rigs. This was, however, not very profitable and during 1977-1978 the trawlers were altered and adjusted for fishing saithe, cod and haddock in Faroese waters. The vessels were equipped with new gear and other equipment. Engine, winch and equipment for the navigating bridge were replaced principally by Norwegian equipment. Except for the fact that 4 of the trawlers are equipped with bigger winches (to be able to fish in deep waters) the 8 trawlers are identical. The gears used are mainly from the same producers and the vessels are similar with respect to construction. However, improvements have been carried out when needed (e.g. winch and engines). Engine power is more than 1000 HP . Total length is about $37-38 \mathrm{~m}$. Since 1985, the mesh size in the trawl is mainly 135 mm (occasionally 145 mm ). Loading capacity is approximately 2000 boxes of fish corresponding to $c a .100$ tons catch per vessel. The trawlers have conducted demersal fisheries around Faeroe Islands since 1977 when the 200 nm boundary was introduced. The Cuba-trawlers started as single trawlers. However, since 1983 the trawlers have operated as pair-trawlers to reduce costs (meaning a reduction of $c a .45 \%$ with respect to fuel and $c a .15 \%$ with respect to fishing gear). The catch is stored on ice on board the trawlers and landed as fresh fish.

The data on which the tuning series are based originate from all available logbooks from the Cuba-trawlers since 1985. The data are stored in the database at the Faroese Fisheries Laboratory in Torshavn, and they are corrected and quality controlled.

The effort obtained from the logbooks is estimated as the number of fishing (trawling) hours, which is the time from when the trawl meets the bottom and until hauling starts. It is not possible to get effort as fishing days because the logbooks do not tell when the trip ends (day and time).

## References

ICES C.M. 1993/Assess:18.

Jákupsstovu, S. March 1999. The Fisheries in Faroese Waters. Fleets, activities, distribution and potential conflicts of interest with an offsshore oil industry.

Steingrund, P. April 2003. Correction of the maturity stages from Faroese spring groundfish survey. WD 14, NWWG 2003.

Table 2.5.1.1 Saithe in the Faroes (Division Vb). Nominal catches (tonnes) by countries, 1989-2002, as officially reported to ICES.

| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Denmark |  |  |  |  |  |  |  |
| Faroe Islands | - | 2 | - | - | - | - | - |
| France $^{3}$ | 43,624 | 59,821 | 53,321 | 35,979 | 32,719 | 32,406 | 26,918 |
| German Dem.Rep. $_{\text {German Fed. Rep. }}^{\text {Netherlands }}$ | - | - | - | 120 | 75 | 19 | 10 |
| Norway | 9 | - | - | 5 | 2 | 1 | 41 |
| UK (Eng. \& W.) | 20 | 15 | 32 |  | - | - | - |
| UK (Scotland) | 51 | 67 | 65 | - | 32 |  |  |
| USSR/Russia ${ }^{2}$ | - | 46 | 103 | 85 | 279 | 156 | 10 |
|  | 9 | 33 | 79 | 74 | 425 | 151 | 21 |
|  | - | 30 | - | 12 | - | - | 200 |


| Total | 43,735 | 60,014 | 53,605 | 36,373 | 33,532 | 33,171 | 27,200 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Working Group estimate ${ }^{4,5}$ | 44,477 | 61,628 | 54,858 | 36,487 | 33,543 | 33,182 | 27,209 |
|  |  |  |  |  |  |  |  |
| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | $2002^{1}$ |


| Estonia | - | 16 | - | - | - | - |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Faroe Islands | 19,297 | 21,721 | 25,995 | 32,439 |  |  |  |
| France | 12 | 9 | 17 | - | 273 | 943 | 705 |
| Germany | 3 | 5 | - | 100 | 230 | 667 | 422 |
| Greenland | - | - | - | - |  |  |  |
| Irland |  |  |  |  |  | 5 |  |
| Norway | 16 | 67 | 53 | 160 | 97 | 80 | 136 |
| Russia | 53 | 28 | - | - | 20 | 1 | 10 |
| UK (E/W/NI) | 580 | 460 | 337 | 441 | 534 | 708 |  |
| UK (Scotland) |  |  |  |  |  |  | 618 |
| United Kingdom | 19,979 | 22,306 | 26,421 | 33,207 | 1,186 | 2,484 | 1,891 |
| Total | 20,029 | 22,306 | 26,421 | 33,207 | 39,045 | 51,795 | 56,759 |
| Working Group estimate$4,5,6$ |  |  |  |  |  |  |  |

${ }^{1}$ Preliminary.
${ }^{2}$ As from 1991.
${ }^{3}$ Quantity unknown 1989-91.
${ }^{4}$ Includes catches from Sub-division Vb2 and Division IIa in Faroese waters.
${ }^{5}$ Includes French, Greenlandic, Russian catches from Division Vb, as reported to the Faroese coastal guard service.
${ }^{6}$ Includes Faroese, French, Greenlandic catches from Division Vb, as reported to the Faroese coastal guard service.

Table 2.5.1.2 Saithe in the Faroes (Division Vb). Total Faroese landings (rightmost column) and the contribution (\%) by each fleet category. Averages for 1985-2002 are given at the bottom.

| Year | Open boats | $\begin{aligned} & \text { Long- } \\ & \text { liners } \\ & \text { <100 GRT } \end{aligned}$ | $\begin{gathered} \text { Single } \\ \text { trawl } \\ <400 \mathrm{HP} \end{gathered}$ | $\begin{aligned} & \text { Gill- } \\ & \text { nets } \end{aligned}$ | Jiggers | Single trawl 4001000 HP | $\begin{gathered} \text { Single } \\ \text { trawl } \\ >1000 \mathrm{HP} \\ \hline \end{gathered}$ | Pair trawl <1000 HP | Pair trawl >1000HP | $\begin{aligned} & \text { Long- } \\ & \text { liners } \\ & >100 \text { GRT } \\ & \hline \end{aligned}$ | Industrial trawlers | Others | Total round weight (tonnes) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 0.2 | 0.1 | 0.1 | 0.0 | 2.6 | 6.6 | 33.7 | 28.2 | 28.2 | 0.1 | 0.2 | 0.2 | 42598 |
| 1986 | 0.3 | 0.2 | 0.1 | 0.1 | 3.6 | 2.8 | 27.3 | 27.5 | 36.5 | 0.1 | 0.7 | 0.9 | 40107 |
| 1987 | 0.7 | 0.1 | 0.3 | 0.4 | 5.6 | 4.1 | 20.4 | 22.8 | 44.2 | 0.1 | 1.1 | 0.0 | 39627 |
| 1988 | 0.4 | 0.3 | 0.1 | 0.3 | 6.5 | 6.8 | 20.8 | 19.6 | 43.6 | 0.1 | 1.3 | 0.1 | 43940 |
| 1989 | 0.9 | 0.1 | 0.3 | 0.2 | 9.3 | 5.4 | 17.7 | 23.5 | 41.1 | 0.1 | 1.3 | 0.0 | 44547 |
| 1990 | 0.6 | 0.2 | 0.2 | 0.2 | 7.4 | 3.9 | 19.6 | 24.0 | 42.8 | 0.2 | 0.9 | 0.0 | 60740 |
| 1991 | 0.6 | 0.1 | 0.1 | 0.6 | 9.8 | 1.3 | 13.9 | 26.5 | 46.2 | 0.1 | 0.8 | 0.0 | 54290 |
| 1992 | 0.4 | 0.4 | 0.0 | 0.0 | 10.5 | 0.5 | 7.1 | 24.4 | 55.6 | 0.1 | 1.0 | 0.0 | 34934 |
| 1993 | 0.6 | 0.2 | 0.1 | 0.0 | 9.3 | 0.6 | 6.5 | 21.4 | 60.6 | 0.1 | 0.7 | 0.0 | 32313 |
| 1994 | 0.4 | 0.4 | 0.1 | 0.0 | 12.6 | 1.1 | 6.8 | 18.5 | 59.1 | 0.2 | 0.7 | 0.0 | 32405 |
| 1995 | 0.2 | 0.1 | 0.4 | 0.0 | 9.6 | 0.9 | 9.9 | 17.7 | 60.9 | 0.3 | 0.0 | 0.0 | 26915 |
| 1996 | 0.0 | 0.0 | 0.1 | 0.0 | 9.2 | 1.2 | 6.8 | 23.7 | 58.6 | 0.2 | 0.0 | 0.0 | 19262 |
| 1997 | 0.0 | 0.1 | 0.1 | 0.0 | 8.9 | 2.5 | 10.7 | 17.8 | 58.9 | 0.4 | 0.4 | 0.0 | 21713 |
| 1998 | 0.1 | 0.4 | 0.1 | 0.0 | 8.1 | 2.8 | 13.8 | 16.5 | 57.6 | 0.3 | 0.4 | 0.0 | 25993 |
| 1999 | 0.0 | 0.1 | 0.1 | 0.0 | 5.7 | 1.2 | 12.6 | 18.5 | 60.0 | 0.2 | 1.6 | 0.0 | 33057 |
| 2000 | 0.1 | 0.1 | 0.2 | 0.0 | 3.7 | 0.3 | 15.0 | 17.5 | 62.3 | 0.1 | 0.7 | 0.0 | 37450 |
| 2001 | 0.1 | 0.1 | 0.1 | 0.0 | 2.8 | 0.3 | 20.2 | 16.5 | 58.8 | 0.2 | 0.8 | 0.1 | 49395 |
| 2002 | 0.1 | 0.2 | 0.1 | 0.0 | 1.4 | 0.1 | 23.9 | 9.5 | 54.8 | 0.1 | 0.0 | 0.0 | 53698 |
| Average | 0.3 | 0.2 | 0.1 | 0.1 | 7.0 | 2.3 | 15.9 | 20.8 | 51.7 | 0.2 | 0.7 | 0.1 | 38499 |

Table 2.5.2.1 Saithe in the Faroes (Division Vb). Catch in number-at-age by fleet categories (calculated from gutted weigths).

| Age | Jiggers | ST $>1000$ Hk | PT<1000 Hk | PT>1000Hk | Others | Tot. Faroe | Foreign | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | 10 | 97 | 58 | 133 | 1 | 307 | 13 | 320 |
| 4 | 107 | 2977 | 1181 | 3976 | 39 | 8505 | 398 | 8903 |
| 5 | 75 | 1199 | 723 | 3964 | 35 | 6160 | 160 | 6320 |
| 6 | 144 | 1737 | 1039 | 6888 | 63 | 10141 | 232 | 10373 |
| 7 | 22 | 264 | 46 | 520 | 4 | 879 | 35 | 914 |
| 8 | 33 | 429 | 113 | 682 | 6 | 1300 | 57 | 1357 |
| 9 | 12 | 177 | 20 | 246 | 2 | 469 | 24 | 493 |
| 10 | 8 | 151 | 18 | 177 | 2 | 364 | 20 | 384 |
| 11 | 1 | 4 | 3 | 25 | 0 | 34 | 1 | 35 |
| 12 | 1 | 10 | 2 | 23 | 0 | 37 | 1 | 38 |
| 13 | 0 | 0 | 0 | 7 | 0 | 8 | 0 | 8 |
| 14 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 1 |
| 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total No. | 414 | 7045 | 3204 | 16639 | 152 | 28205 | 941 | 29146 |
| Catch, $t$. | 871 | 14151 | 5628 | 32465 | 294 | 54868 | 1891 | 56759 |

Notes:
Numbers in $1000^{\prime}$
Catch, round weight in tonnes
ST- single trawlers and PT- pair trawlers
Others includes longliners, small single trawlers, industrial trawlers and catches not otherwise accounted for

Table 2.5.2.2 Saithe in the Faroes (Division Vb). Catch numbers-at-age (Thousands).


Table 2.5.3.1 Saithe in the Faroes (Division Vb ). Catch weights-at-age (kg).


Table 2.5.4.1 Saithe in the Faroes (Division Vb). Proportion mature-at-age.


Table 2.5.5.1 Saithe in the Faroes (Division Vb). Effort (hours) and catch in number-at-age for commercial Cuba Logbook pair trawlers.

| Faroe Saithe (ICES Div. Vb) age3+5-11102 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Cuba Logbook age3 |  |  |  |  |  |  |  |
| 19852002 |  |  |  |  |  |  |  |
| 1101 |  |  |  |  |  |  |  |
| 33 |  |  |  |  |  |  |  |
| 320030 |  |  |  |  |  |  |  |
| $3565 \quad 63$ |  |  |  |  |  |  |  |
| 6470117 |  |  |  |  |  |  |  |
| 697565 |  |  |  |  |  |  |  |
| $6635 \quad 51$ |  |  |  |  |  |  |  |
| 789223 |  |  |  |  |  |  |  |
| 776234 |  |  |  |  |  |  |  |
| 788411 |  |  |  |  |  |  |  |
| 774781 |  |  |  |  |  |  |  |
| 729078 |  |  |  |  |  |  |  |
| $8067 \quad 70$ |  |  |  |  |  |  |  |
| 534532 |  |  |  |  |  |  |  |
| 730679 |  |  |  |  |  |  |  |
| 830219 |  |  |  |  |  |  |  |
| 1163957 |  |  |  |  |  |  |  |
| 9845138 |  |  |  |  |  |  |  |
| 11007231 |  |  |  |  |  |  |  |
| 964841 |  |  |  |  |  |  |  |
| Cuba Logbook age5-11 |  |  |  |  |  |  |  |
| 19852002 |  |  |  |  |  |  |  |
| 1101 |  |  |  |  |  |  |  |
| 511 |  |  |  |  |  |  |  |
| 3200 | 386 | 81 | 100 | 13 | 5 | 1 | 2 |
| 3565 | 332 | 238 | 66 | 70 | 17 | 6 | 3 |
| 6470 | 391 | 311 | 116 | 61 | 32 | 7 | 5 |
| 6976 | 865 | 304 | 123 | 51 | 35 | 8 | 1 |
| 6635 | 497 | 628 | 72 | 42 | 14 | 8 | 5 |
| 7892 | 737 | 713 | 392 | 36 | 10 | 4 | 4 |
| 7762 | 491 | 397 | 250 | 123 | 15 | 14 | 10 |
| 7884 | 271 | 229 | 115 | 67 | 58 | 11 | 6 |
| 7747 | 514 | 185 | 99 | 59 | 43 | 25 | 6 |
| 7290 | 356 | 334 | 107 | 75 | 49 | 45 | 21 |
| 8067 | 852 | 349 | 215 | 72 | 36 | 28 | 20 |
| 5345 | 169 | 174 | 213 | 163 | 79 | 20 | 11 |
| 7306 | 406 | 298 | 229 | 104 | 24 | 12 | 4 |
| 8302 | 306 | 532 | 226 | 112 | 63 | 12 | 4 |
| 11639 | 668 | 608 | 988 | 240 | 105 | 35 | 6 |
| 9845 | 301 | 905 | 463 | 571 | 70 | 39 | 12 |
| 11007 | 1865 | 439 | 700 | 282 | 234 | 48 | 11 |
| 9648 | 1225 | 2129 | 161 | 210 | 76 | 54 | 8 |

Table 2.5.5.2

| 2/05/2003 16:59 |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Extended Survivors Analysis |  |  |  |  |  |  |  |  |  |  |
| FAROE SAITHE (ICES Division Vb) SAI_IND |  |  |  |  |  |  |  |  |  |  |
| CPUE data from file D:\Stovnsmeting\Ices2003\Koyringar\Xsa\saithetun. DAT |  |  |  |  |  |  |  |  |  |  |
| Catch data for 42 years. 1961 to 2002. Ages 3 to 12. |  |  |  |  |  |  |  |  |  |  |
| Fleet |  |  | First | Last | First | Last | Alpha | eta |  |  |
|  |  |  | year | year | age | age |  |  |  |  |
| Cuba Logbook age3 |  |  | 1985 | 2002 | 3 | 3 | . 000 | . 000 |  |  |
| Cuba Logbook age5-11 |  |  | 1985 | 2002 | 5 | 11 | . 000 | . 000 |  |  |
| Time-series weights : |  |  |  |  |  |  |  |  |  |  |
| Tapered time weighting applied |  |  |  |  |  |  |  |  |  |  |
| Power = 3 over 20 years |  |  |  |  |  |  |  |  |  |  |
| Catchability analysis : |  |  |  |  |  |  |  |  |  |  |
| Catchability dependent on stock size for ages < 5 |  |  |  |  |  |  |  |  |  |  |
| Regression type $=C$ <br> Minimum of 5 points used for regression <br> Survivor estimates shrunk to the population mean for ages |  |  |  |  |  |  |  |  |  |  |
| Catchability independent of age for ages >= 9 |  |  |  |  |  |  |  |  |  |  |
| Terminal population estimation : |  |  |  |  |  |  |  |  |  |  |
| Survivor estimates shrunk towards the mean F of the final 5 years or the 3 oldest ages. |  |  |  |  |  |  |  |  |  |  |
| S.E. of the mean to which the estimates are shrunk = . 500 |  |  |  |  |  |  |  |  |  |  |
| Minimum standard error for population estimates derived from each fleet $=$. 300 |  |  |  |  |  |  |  |  |  |  |
| Prior weighting not applied |  |  |  |  |  |  |  |  |  |  |
| Tuning converged after 18 iterations |  |  |  |  |  |  |  |  |  |  |
| Regression weights |  |  |  |  |  |  |  |  |  |  |
|  | . 751 | . 820 | . 877 | . 921 | . 954 | . 976 | . 990 | . 997 | 1.000 | 1.000 |
| Fishing mortalities |  |  |  |  |  |  |  |  |  |  |
| Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| 3 | . 063 | . 047 | . 011 | . 014 | . 011 | . 013 | . 004 | . 014 | . 012 | . 012 |
| 4 | . 205 | . 274 | . 090 | . 039 | . 050 | . 069 | . 069 | . 049 | . 054 | . 125 |
| 5 | . 553 | . 333 | . 411 | . 139 | . 114 | . 157 | . 175 | . 220 | . 203 | . 193 |
| 6 | . 604 | . 607 | . 404 | . 301 | . 328 | . 237 | . 320 | . 402 | . 571 | . 413 |
| 7 | . 520 | . 606 | . 707 | . 483 | . 502 | . 463 | . 489 | . 515 | . 710 | . 509 |
| 8 | . 417 | . 664 | . 636 | . 812 | . 526 | . 523 | . 648 | . 714 | . 841 | . 627 |
| 9 | . 492 | . 502 | . 809 | . 536 | . 663 | . 690 | . 629 | . 551 | . 973 | . 726 |
| 10 | . 500 | . 744 | . 688 | . 666 | . 628 | . 768 | . 657 | . 741 | 1.409 | . 864 |
| 11 | . 625 | . 663 | . 621 | . 516 | . 571 | . 656 | . 548 | . 572 | 1.047 | . 600 |

## Table 2.5.5.2. (Continued)

|  |  |  | AGE |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
| 1993 | 2.38E+04 | $1.56 \mathrm{E}+04$ | 1.22E+04 | $4.06 \mathrm{E}+03$ | $2.34 \mathrm{E}+03$ | $1.59 \mathrm{E}+03$ | $1.27 \mathrm{E}+03$ | $6.89 \mathrm{E}+02$ | $1.28 \mathrm{E}+02$ |
| 1994 | 1.67E+04 | $1.83 \mathrm{E}+04$ | $1.04 \mathrm{E}+04$ | $5.75 \mathrm{E}+03$ | $1.81 \mathrm{E}+03$ | $1.14 \mathrm{E}+03$ | $8.60 \mathrm{E}+02$ | $6.38 \mathrm{E}+02$ | $3.42 \mathrm{E}+02$ |
| 1995 | $3.91 \mathrm{E}+04$ | $1.31 \mathrm{E}+04$ | $1.14 \mathrm{E}+04$ | $6.10 \mathrm{E}+03$ | $2.57 \mathrm{E}+03$ | $8.11 \mathrm{E}+02$ | $4.80 \mathrm{E}+02$ | $4.27 \mathrm{E}+02$ | $2.48 \mathrm{E}+02$ |
| 1996 | 2.34E+04 | $3.17 \mathrm{E}+04$ | 9.79E+03 | $6.17 \mathrm{E}+03$ | $3.34 \mathrm{E}+03$ | $1.04 \mathrm{E}+03$ | $3.51 \mathrm{E}+02$ | $1.75 \mathrm{E}+02$ | 1.75E+02 |
| 1997 | $3.44 \mathrm{E}+04$ | $1.89 \mathrm{E}+04$ | $2.49 \mathrm{E}+04$ | $6.98 \mathrm{E}+03$ | $3.74 \mathrm{E}+03$ | $1.69 \mathrm{E}+03$ | $3.76 \mathrm{E}+02$ | $1.68 \mathrm{E}+02$ | $7.37 \mathrm{E}+01$ |
| 1998 | 1.35E+04 | $2.78 \mathrm{E}+04$ | $1.47 \mathrm{E}+04$ | $1.82 \mathrm{E}+04$ | $4.11 \mathrm{E}+03$ | $1.85 \mathrm{E}+03$ | $8.16 \mathrm{E}+02$ | $1.59 \mathrm{E}+02$ | $7.35 \mathrm{E}+01$ |
| 1999 | $7.98 \mathrm{E}+04$ | $1.09 \mathrm{E}+04$ | $2.13 \mathrm{E}+04$ | $1.03 \mathrm{E}+04$ | $1.18 \mathrm{E}+04$ | $2.12 \mathrm{E}+03$ | $9.00 \mathrm{E}+02$ | $3.35 \mathrm{E}+02$ | $6.03 \mathrm{E}+01$ |
| 2000 | $6.35 E+04$ | $6.50 \mathrm{E}+04$ | $8.33 \mathrm{E}+03$ | $1.46 \mathrm{E}+04$ | $6.11 \mathrm{E}+03$ | $5.91 \mathrm{E}+03$ | $9.08 \mathrm{E}+02$ | $3.93 \mathrm{E}+02$ | 1.42E+02 |
| 2001 | $1.04 \mathrm{E}+05$ | $5.13 \mathrm{E}+04$ | $5.07 \mathrm{E}+04$ | $5.47 \mathrm{E}+03$ | $8.00 \mathrm{E}+03$ | $2.99 \mathrm{E}+03$ | $2.37 \mathrm{E}+03$ | $4.28 \mathrm{E}+02$ | 1.53E+02 |
| 2002 | $2.95 E+04$ | 8.40E+04 | $3.97 \mathrm{E}+04$ | $3.39 \mathrm{E}+04$ | $2.53 \mathrm{E}+03$ | $3.22 \mathrm{E}+03$ | $1.06 \mathrm{E}+03$ | $7.33 \mathrm{E}+02$ | $8.57 \mathrm{E}+01$ |

Estimated population abundance at 1st Jan 2003
$0.00 \mathrm{E}+00 \quad 2.39 \mathrm{E}+04 \quad 6.07 \mathrm{E}+04 \quad 2.68 \mathrm{E}+04 \quad 1.83 \mathrm{E}+04 \quad 1.25 \mathrm{E}+03 \quad 1.41 \mathrm{E}+03 \quad 4.18 \mathrm{E}+02 \quad 2.53 \mathrm{E}+02$
Taper weighted geometric mean of the VPA populations:
$\begin{array}{lllllllll}3.35 \mathrm{E}+04 & 2.68 \mathrm{E}+04 & 1.72 \mathrm{E}+04 & 9.37 \mathrm{E}+03 & 4.14 \mathrm{E}+03 & 1.96 \mathrm{E}+03 & 8.15 \mathrm{E}+02 & 3.43 \mathrm{E}+02 & 1.28 \mathrm{E}+02\end{array}$
Standard error of the weighted Log(VPA populations) :
$.6269 .6468 \quad .5981 \quad .6308 \quad .5753$. 5577 . $5404 \quad .5265$. 5291

Log-catchability residuals.
Fleet : Cuba Logbook age3

| Age | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | . 89 | . 71 | . 99 | . 15 | . 33 | -. 68 | -. 28 | -1.63 |  |  |
| 4 | No data | for t | this fle | t at t | this age |  |  |  |  |  |
| 5 | No data | for t | this fle | $t$ at t | this age |  |  |  |  |  |
| 6 | No data | for t | this fle | $t$ at $t$ | this age |  |  |  |  |  |
| 7 | No data | for t | this fle | $t$ at | this age |  |  |  |  |  |
| 8 | No data | for t | this fle | $t$ at | this age |  |  |  |  |  |
| 9 | No data | for t | this fle | $t$ at | this age |  |  |  |  |  |
| 10 | No data | for t | this fle | $t$ at t | this age |  |  |  |  |  |
| 11 | No data | for t | this fle | $t$ at $t$ | this age |  |  |  |  |  |
| Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| 3 | . 97 | 1.34 | . 18 | . 18 | . 61 | -. 59 | -1.33 | . 36 | . 42 | -. 52 |
| 4 | No data | for t | this fle | t at t | this age |  |  |  |  |  |
| 5 | No data | for t | this fle | $t$ at | this age |  |  |  |  |  |
| 6 | No data | for t | this fle | $t$ at | this age |  |  |  |  |  |
| 7 | No data | for t | this fle | $t$ at | this age |  |  |  |  |  |
| 8 | No data | for t | this fle | $t$ at t | this age |  |  |  |  |  |
| 9 | No data | for t | this fle | $t$ at t | this age |  |  |  |  |  |
| 10 | No data | for t | this fle | $t$ at | this age |  |  |  |  |  |
| 11 | No data | for t | this fle | t at t | this age |  |  |  |  |  |

Regression statistics :
Ages with $q$ dependent on year class strength

| Age Slope | t-value | Intercept | RSquare | No Pts | Reg s.e Mean Log q |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3 | 1.38 | -.898 | 17.14 | .36 | 18 | .87 | -15.30 |

Fleet : Cuba Logbook age5-11

| Age | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: | ---: | ---: |
| 3 | No data for this fleet at this age |  |  |  |  |  |  |  |
| 4 | No data for this fleet at this age |  |  |  |  |  |  |  |
| 5 | .85 | .71 | .37 | -.02 | -.37 | .21 | .32 | .17 |
| 6 | -.17 | .55 | .23 | .24 | -.20 | -.04 | .02 | -.04 |
| 7 | -.05 | -.15 | -.15 | -.21 | -.64 | -.25 | -.31 | -.51 |
| 8 | -.58 | .23 | -.28 | -.42 | -.57 | -1.03 | -.56 | -.90 |
| 9 | -.81 | .41 | -.22 | -.10 | -.93 | -1.57 | -1.09 | -.48 |
| 10 | -1.80 | -.05 | .01 | -.84 | -.74 | -1.84 | -.47 | -.48 |
| 11 | -.33 | -.12 | .04 | -.95 | -.43 | -.97 | -.16 | -.28 |

## Table 2.5.5.2. (Continued)

| Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: | ---: |
| 3 | No data for this fleet at this age |  |  |  |  |  |  |  |  |  |
| 4 | No data for this fleet at this age |  |  |  |  |  |  |  |  |  |
| 5 | .51 | .26 | .98 | -.20 | -.59 | -.45 | -.37 | -.04 | -.14 | -.19 |
| 6 | .08 | .38 | .17 | -.17 | -.05 | -.60 | -.20 | .05 | .27 | .09 |
| 7 | -.27 | .16 | .45 | .50 | .15 | -.10 | -.01 | .07 | .19 | -.09 |
| 8 | -.58 | .17 | .36 | 1.41 | .04 | -.11 | .24 | .27 | .19 | -.14 |
| 9 | -.63 | -.04 | .26 | 1.66 | .14 | .21 | .26 | -.02 | .30 | .01 |
| 10 | -.55 | .28 | .08 | 1.04 | .24 | .23 | .16 | .32 | .59 | .09 |
| 11 | -.25 | .10 | .25 | .37 | -.06 | -.15 | .07 | .08 | .01 | .21 |

Mean log-catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q | -12.2750 | -11.7431 | -11.5094 | -11.3835 | -11.3869 | -11.3869 |
| S.E (Log q) | .4379 | .2589 | .3005 | .5924 | .7078 | .6296 |

Regression statistics :
Ages with $q$ independent of year class strength and constant w.r.t. time.
Age Slope t-value Intercept RSquare No Pts Reg s.e Mean Q

| 5 | 1.42 | -1.407 | 13.34 | .52 | 18 | .60 | -12.28 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 6 | 1.19 | -1.339 | 12.24 | .83 | 18 | .30 | -11.74 |
| 7 | 1.03 | -.186 | 11.61 | .78 | 18 | .32 | -11.51 |
| 8 | 1.32 | -.749 | 12.61 | .35 | 18 | .80 | -11.38 |
| 9 | 1.69 | -1.035 | 14.60 | .18 | 18 | 1.19 | -11.39 |
| 10 | 1.35 | -.700 | 13.27 | .29 | 18 | .87 | -11.36 |
| 11 | .89 | .627 | 10.74 | .78 | 18 | .30 | -11.44 |

Terminal year survivor and $F$ summaries :
Age 3 Catchability dependent on age and year class strength

Year class $=1999$

| Fleet | Estimated Survivors | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled Weights | $\begin{gathered} \text { Estimated } \\ \mathrm{F} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Cuba Logbook age3 | 14172. | . 922 | . 000 | . 00 | 1 | . 154 | . 020 |
| Cuba Logbook age5-11 | 1. | . 000 | . 000 | . 00 | 0 | . 000 | . 000 |
| $P$ shrinkage mean | 26768. | . 65 |  |  |  | . 316 | . 011 |
| F shrinkage mean | 25989. | . 50 |  |  |  | . 530 | . 011 |

Weighted prediction :

| Survivors | Int | Ext | $N$ | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 23895. | .36 | .40 | 3 | 1.105 | .012 |

Age 4 Catchability dependent on age and year class strength
Year class $=1998$

| Fleet | Estimated Survivors | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | $\begin{gathered} \text { Var } \\ \text { Ratio } \end{gathered}$ | N | Scaled Weights | $\begin{gathered} \text { Estimated } \\ \mathrm{F} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Cuba Logbook age3 | 92161. | . 988 | . 000 | .00 | 1 | . 116 | . 084 |
| Cuba Logbook age5-11 | 1. | . 000 | .000 | .00 | 0 | .000 | . 000 |
| $P$ shrinkage mean | 17188. | . 60 |  |  |  | . 364 | . 384 |
| F shrinkage mean | 133719. | . 50 |  |  |  | . 520 | . 058 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | S.e |  | Ratio |  |
| 60739. | .36 | .74 | 3 | 2.070 | .125 |

## Table 2.5.5.2. (Continued)

| Year class $=1997$ |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Cuba Logbook age3 | 38268. | . 943 | . 000 | . 00 | 1 | . 098 | . 139 |
| Cuba Logbook age5-11 | 22207. | . 456 | . 000 | . 00 | 1 | . 449 | . 229 |
| F shrinkage mean | 29957. | . 50 |  |  |  | . 453 | . 175 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 26824. | .32 | .15 | 3 | .456 | .193 |

Age 6 Catchability constant w.r.t. time and dependent on age
Year class $=1996$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :--- | ---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Curvivors | S.e | S.e | Ratio |  | Weights | $F$ |  |
| Cuba Logbook age3 | 4870. | .919 | .000 | .00 | 1 | .042 | 1.074 |
| Cuba Logbook age5-11 | 18944. | .252 | .102 | .40 | 2 | .684 | .402 |
|  |  |  |  |  |  |  |  |
| F shrinkage mean | 20689. | .50 |  |  |  | .275 | .374 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 18342. | .22 | .17 | 4 | .762 | .413 |

Age 7 Catchability constant w.r.t. time and dependent on age
Year class $=1995$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Cuba Logbook age3 | 692. | . 994 | . 000 | . 00 | 1 | . 016 | . 787 |
| Cuba Logbook age5-11 | 1297. | . 206 | . 118 | . 57 | 3 | . 725 | .493 |
| F shrinkage mean | 1155. | . 50 |  |  |  | . 258 | . 540 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | S.e |  | Ratio |  |
| 1246. | .20 | .09 | 5 | .432 | .509 |

Age 8 Catchability constant w.r.t. time and dependent on age
Year class $=1994$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Cuba Logbook age3 | 2594. | . 943 | . 000 | . 00 | 1 | . 014 | 388 |
| Cuba Logbook age5-11 | 1443. | . 205 | .103 | . 51 | 4 | . 619 | . 616 |
| F shrinkage mean | 1324. | . 50 |  |  |  | . 367 | . 656 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 1409 | .22 | .08 | 6 | .336 | .627 |

## Table 2.5.5.2. (Continued)

Age 9 Catchability constant w.r.t. time and dependent on age

| Year class $=1993$ |  |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| Cuba Logbook age3 | Survivors | S.e | S.e | Ratio | Weights | F |  |
| Cuba Logbook age5-11 | 503. | .946 | .000 | .00 | 1 | .010 | .635 |
|  | 402. | .223 | .090 | .41 | 5 | .499 | .746 |
| F shrinkage mean | 433. | .50 |  |  |  | .491 | .708 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 418. | .27 | .06 | 7 | .210 | .726 |

Age 10 Catchability constant w.r.t. time and age (fixed at the value for age) 9

| Year class $=1992$ |  |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| Cuba Logbook age3 | Survivors | S.e | s.e | Ratio | Weights | F |  |
| Cuba Logbook age5-11 | 302. | .973 | .000 | .00 | 1 | .005 | .765 |
|  | 242. | .285 | .140 | .49 | 6 | .385 | .890 |
| F shrinkage mean | 260. | .50 |  |  |  | .610 | .849 |


| Weighted prediction : |  |  |  |  |  |
| :--- | ---: | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |
| at end of year | s.e | s.e |  | Ratio |  |
| 253. | .32 | .08 | 8 | .235 | .864 |

Age 11 Catchability constant w.r.t. time and age (fixed at the value for age) 9
Year class = 1991

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :--- | ---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Cuba Logbook age3 | Survors | S.e | S.e | Ratio | Weights | F |  |
| Cuba Logbook age5-11 | 147. | 1.018 | .000 | .00 | 1 | .001 | .195 |
|  | 47. | .280 | .054 | .19 | 7 | .593 | .514 |
| F shrinkage mean | 29. | .50 |  |  |  | .406 | .746 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year | S.e | S.e |  | Ratio |  |
| 39. | .26 | .12 | 9 | .457 | .600 |



Table 2.5.5.4
Saithe in the Faroes (Division Vb). Stock number-at-age (start of year) (Thousands).

Run title : FAROE SAITHE (ICES Division Vb)
SAI IND
At 2/05/2003 19:19
Traditional vpa using screen input for terminal $F$ with backwards extension


|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 23548 | 16587 |  | 38708 | 23153 | 34050 | 13367 | 79201 | 63192 | 79999 | 22756 | 0 | 23737 | 27288 |
|  | 4 | 15390 | 18092 |  | 12957 | 31332 | 18688 | 27567 | 10797 | 64554 | 51004 | 64482 | 18342 | 17554 | 20294 |
|  | 5 | 12033 | 10250 |  | 11251 | 9689 | 24671 | 14550 | 21046 | 8249 | 50296 | 39545 | 44773 | 11404 | 12986 |
|  | 6 | 4000 | 5655 |  | 6000 | 6100 | 6900 | 17999 | 10170 | 14443 | 5417 | 33583 | 26686 | 6866 | 7837 |
|  | 7 | 2300 | 1786 |  | 2512 | 3265 | 3692 | 4062 | 11610 | 6035 | 7902 | 2506 | 18189 | 3778 | 4312 |
|  | 8 | 1516 | 1115 |  | 795 | 1006 | 1637 | 1827 | 2089 | 5820 | 2947 | 3183 | 1233 | 2080 | 2427 |
|  | 9 | 1212 | 800 |  | 466 | 343 | 359 | 782 | 884 | 893 | 2330 | 1041 | 1393 | 1131 | 1401 |
|  | 10 | 674 | 591 |  | 380 | 167 | 162 | 147 | 311 | 384 | 420 | 722 | 412 | 635 | 874 |
|  | 11 | 153 | 332 |  | 214 | 140 | 68 | 70 | 52 | 125 | 148 | 85 | 249 | 359 | 550 |
|  | +gp | 147 | 109 |  | 240 | 179 | 234 | 159 | 155 | 56 | 92 | 114 | 89 |  |  |
| 0 | TOTAL | 60974 | 55316 |  | 73522 | 75374 | 90462 | 80528 | 136314 | 163749 | 200555 | 168016 | 111366 |  |  |

Table 2.5.5.5
Saithe in the Faroes (Division Vb). Summary table.

Run title : FAROE SAITHE (ICES Division Vb)
SAI_IND
At 2/05/2003 19:19

Table 16 Summary (without SOP correction)
Traditional vpa using screen input for terminal $F$ with backwards extension

|  | RECRUITS <br> Age 3 | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | FBAR | 4-8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 9032 | 122159 | 84047 | 9592 | . 1141 |  | . 0907 |
| 1962 | 13619 | 126558 | 85825 | 10454 | . 1218 |  | . 1080 |
| 1963 | 22363 | 158306 | 100859 | 12693 | . 1258 |  | . 0993 |
| 1964 | 16181 | 160324 | 98419 | 21893 | . 2224 |  | . 2000 |
| 1965 | 22750 | 174701 | 107272 | 22181 | . 2068 |  | . 1821 |
| 1966 | 21787 | 184036 | 108806 | 25563 | . 2349 |  | . 2020 |
| 1967 | 26822 | 181502 | 104636 | 21319 | . 2037 |  | . 1653 |
| 1968 | 21451 | 189683 | 116011 | 20387 | . 1757 |  | . 1345 |
| 1969 | 40612 | 214702 | 123787 | 27437 | . 2216 |  | . 1783 |
| 1970 | 34010 | 224052 | 129102 | 29110 | . 2255 |  | . 1828 |
| 1971 | 37084 | 227929 | 139397 | 32706 | . 2346 |  | . 1764 |
| 1972 | 33414 | 236417 | 147387 | 42663 | . 2895 |  | . 2318 |
| 1973 | 23106 | 209953 | 136561 | 57431 | . 4206 |  | . 3314 |
| 1974 | 18771 | 203579 | 137545 | 47188 | . 3431 |  | . 2804 |
| 1975 | 16196 | 187008 | 137809 | 41576 | . 3017 |  | . 3120 |
| 1976 | 18780 | 169263 | 121855 | 33065 | . 2713 |  | . 2818 |
| 1977 | 12842 | 155790 | 113860 | 34835 | . 3059 |  | . 3509 |
| 1978 | 8357 | 136872 | 95807 | 28138 | . 2937 |  | . 2658 |
| 1979 | 8567 | 112662 | 83398 | 27246 | . 3267 |  | . 2848 |
| 1980 | 12346 | 124362 | 88748 | 25230 | . 2843 |  | . 2331 |
| 1981 | 33021 | 141447 | 76135 | 30103 | . 3954 |  | . 4122 |
| 1982 | 15097 | 149398 | 83124 | 30964 | . 3725 |  | . 3457 |
| 1983 | 40553 | 177824 | 92868 | 39176 | . 4218 |  | . 3916 |
| 1984 | 25707 | 188593 | 96051 | 54665 | . 5691 |  | . 5020 |
| 1985 | 21951 | 188334 | 109372 | 44605 | . 4078 |  | . 4032 |
| 1986 | 61014 | 233026 | 98143 | 41716 | . 4251 |  | . 5021 |
| 1987 | 47827 | 247283 | 94536 | 40020 | . 4233 |  | . 4057 |
| 1988 | 43910 | 256264 | 100171 | 45285 | . 4521 |  | . 4563 |
| 1989 | 28200 | 225255 | 99274 | 44477 | . 4480 |  | . 3696 |
| 1990 | 20449 | 188887 | 93772 | 61628 | . 6572 |  | . 5728 |
| 1991 | 24543 | 147107 | 71052 | 54858 | . 7721 |  | . 7180 |
| 1992 | 19372 | 121718 | 59186 | 36487 | . 6165 |  | . 5356 |
| 1993 | 23548 | 130691 | 61828 | 33543 | . 5425 |  | . 4663 |
| 1994 | 16587 | 124455 | 59908 | 33182 | . 5539 |  | . 5006 |
| 1995 | 38708 | 150564 | 62435 | 27209 | . 4358 |  | . 4537 |
| 1996 | 23153 | 158785 | 68859 | 20029 | . 2909 |  | . 3601 |
| 1997 | 34050 | 180059 | 75916 | 22306 | . 2938 |  | . 3077 |
| 1998 | 13367 | 165206 | 87637 | 26421 | . 3015 |  | . 2914 |
| 1999 | 79201 | 241457 | 103622 | 33207 | . 3205 |  | . 3417 |
| 2000 | 63192 | 293940 | 107199 | 39045 | . 3642 |  | . 3811 |
| 2001 | 79999 | 341019 | 124784 | 51795 | . 4151 |  | . 4757 |
| 2002 | 22756 | 280661 | 122102 | 56759 | . 4648 |  | . 3814 |
| Arith. |  |  |  |  |  |  |  |
| Mean | 28436 | 186472 | 100217 | 34243 | . 3540 |  | . 3301 |
| 0 Units | (Thousands) | (Tonnes) | (Tonnes) | (Tonnes) |  |  |  |

Saithe in the Faroes (Division Vb). Predictions with management table. A) input data based on the run where 1998 year class was reset to approximately the highest observed B) input data based on the predicted 1998 year-class strength from the XSA.

## A)

MFDP version 1a
Run: man9
Time and date: 09:25 07/05/03
Fbar age range: 4-8

2003

| Age | N | M | Mat | PF |  | PM |  | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 29649 | 0.2 | 0.09 |  | 0 |  | 0 | 1.316 | 0.015 | 1.316 |
| 4 | 18342 | 0.2 | 0.21 |  | 0 |  | 0 | 1.552 | 0.090 | 1.552 |
| 5 | 44773 | 0.2 | 0.38 |  | 0 |  | 0 | 1.838 | 0.206 | 1.838 |
| 6 | 26686 | 0.2 | 0.60 |  | 0 |  | 0 | 2.305 | 0.462 | 2.305 |
| 7 | 18189 | 0.2 | 0.88 |  | 0 |  | 0 | 3.055 | 0.578 | 3.055 |
| 8 | 1233 | 0.2 | 0.95 |  | 0 |  | 0 | 3.647 | 0.727 | 3.647 |
| 9 | 1393 | 0.2 | 0.99 |  | 0 |  | 0 | 4.579 | 0.751 | 4.579 |
| 10 | 412 | 0.2 | 1.00 |  | 0 |  | 0 | 5.324 | 1.005 | 5.324 |
| 11 | 249 | 0.2 | 1.00 |  | 0 |  | 0 | 6.549 | 0.782 | 6.549 |
| 12 | 89 | 0.2 | 1.00 |  | 0 |  | 0 | 6.784 | 0.782 | 6.784 |



| 2005 |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N | M | Mat | PF |  | PM | SWt | Sel | CWt |
| 3 | 29649 | 0.2 | 0.11 |  | 0 | 0 | 1.316 | 0.015 | 1.316 |
| 4 |  | 0.2 | 0.22 |  | 0 | 0 | 1.552 | 0.090 | 1.552 |
| 5 |  | 0.2 | 0.40 |  | 0 | 0 | 1.838 | 0.206 | 1.838 |
| 6 |  | 0.2 | 0.65 |  | 0 | 0 | 2.305 | 0.462 | 2.305 |
| 7 |  | 0.2 | 0.87 |  | 0 | 0 | 3.055 | 0.578 | 3.055 |
| 8 |  | 0.2 | 0.95 |  | 0 | 0 | 3.647 | 0.727 | 3.647 |
| 9 |  | 0.2 | 0.99 |  | 0 | 0 | 4.579 | 0.751 | 4.579 |
| 10 |  | 0.2 | 1.00 |  | 0 | 0 | 5.324 | 1.005 | 5.324 |
| 11 |  | 0.2 | 1.00 |  | 0 | 0 | 6.549 | 0.782 | 6.549 |
| 12 |  | 0.2 | 1.00 |  | 0 | 0 | 6.784 | 0.782 | 6.784 |

Input units are thousands and kg - output in tonnes

Table 2.5.6.1. Continued

## B)

MFDP version 1a
Run: man9a
Time and date: 09:33 07/05/03
Fbar age range: 4-8

## 2003

| Age | N | M | Mat | PF | PM |  | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 30520 | 0.2 | 0.09 |  | 0 | 0 | 1.316 | 0.013 | 1.316 |
| 4 | 23895 | 0.2 | 0.21 |  | 0 | 0 | 1.552 | 0.076 | 1.552 |
| 5 | 60739 | 0.2 | 0.38 |  | 0 | 0 | 1.838 | 0.205 | 1.838 |
| 6 | 26824 | 0.2 | 0.60 |  | 0 | 0 | 2.305 | 0.462 | 2.305 |
| 7 | 18342 | 0.2 | 0.88 |  | 0 | 0 | 3.055 | 0.578 | 3.055 |
| 8 | 1246 | 0.2 | 0.95 |  | 0 | 0 | 3.647 | 0.727 | 3.647 |
| 9 | 1409 | 0.2 | 0.99 |  | 0 | 0 | 4.579 | 0.750 | 4.579 |
| 10 | 418 | 0.2 | 1.00 |  | 0 | 0 | 5.324 | 1.005 | 5.324 |
| 11 | 253 | 0.2 | 1.00 |  | 0 | 0 | 6.549 | 0.739 | 6.549 |
| 12 | 90 | 0.2 | 1.00 |  | 0 | 0 | 6.784 | 0.739 | 6.784 |

2004

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 30520 | 0.2 | 0.11 | 0 | 0 | 1.316 | 0.013 | 1.316 |
| 4. |  | 0.2 | 0.22 | 0 | 0 | 1.552 | 0.076 | 1.552 |
| 5. |  | 0.2 | 0.40 | 0 | 0 | 1.838 | 0.205 | 1.838 |
| 6. |  | 0.2 | 0.65 | 0 | 0 | 2.305 | 0.462 | 2.305 |
| 7. |  | 0.2 | 0.87 | 0 | 0 | 3.055 | 0.578 | 3.055 |
| 8. |  | 0.2 | 0.95 | 0 | 0 | 3.647 | 0.727 | 3.647 |
| 9. |  | 0.2 | 0.99 | 0 | 0 | 4.579 | 0.750 | 4.579 |
| 10. |  | 0.2 | 1.00 | 0 | 0 | 5.324 | 1.005 | 5.324 |
| 11. |  | 0.2 | 1.00 | 0 | 0 | 6.549 | 0.739 | 6.549 |
| 12. |  | 0.2 | 1.00 | 0 | 0 | 6.784 | 0.739 | 6.784 |

2005

| Age | N | M | Mat | PF |  | PM |  | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 30520 | 0.2 | 0.11 |  | 0 |  | 0 | 1.316 | 0.013 | 1.316 |
| 4. |  | 0.2 | 0.22 |  | 0 |  | 0 | 1.552 | 0.076 | 1.552 |
| 5. |  | 0.2 | 0.40 |  | 0 |  | 0 | 1.838 | 0.205 | 1.838 |
| 6. |  | 0.2 | 0.65 |  | 0 |  | 0 | 2.305 | 0.462 | 2.305 |
| 7. |  | 0.2 | 0.87 |  | 0 |  | 0 | 3.055 | 0.578 | 3.055 |
| 8. |  | 0.2 | 0.95 |  | 0 |  | 0 | 3.647 | 0.727 | 3.647 |
| 9. |  | 0.2 | 0.99 |  | 0 |  | 0 | 4.579 | 0.750 | 4.579 |
| 10. |  | 0.2 | 1.00 |  | 0 |  | 0 | 5.324 | 1.005 | 5.324 |
| 11. |  | 0.2 | 1.00 |  | 0 |  | 0 | 6.549 | 0.739 | 6.549 |
| 12. |  | 0.2 | 1.00 |  | 0 |  | 0 | 6.784 | 0.739 | 6.784 |

Input units are thousands and kg - output in tonnes year class was reset to approximately the highest observed B) input data based on the predicted 1998 year-class strength from the XSA.

## A)

MFYPR version 2 a
Run: yr2
Index file 3/5/2003
Time and date: 19:46 05/05/03
Fbar age range: 4-8

| Age | M | Mat | PF | PM | SWt | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.2 | 0.11 | 0 | 0 | 1.316 | 0.053 | 1.316 |
| 4 | 0.2 | 0.26 | 0 | 0 | 1.822 | 0.173 | 1.822 |
| 5 | 0.2 | 0.48 | 0 | 0 | 2.444 | 0.294 | 2.444 |
| 6 | 0.2 | 0.74 | 0 | 0 | 3.183 | 0.376 | 3.183 |
| 7 | 0.2 | 0.91 | 0 | 0 | 4.040 | 0.395 | 4.040 |
| 8 | 0.2 | 0.98 | 0 | 0 | 4.956 | 0.412 | 4.956 |
| 9 | 0.2 | 0.99 | 0 | 0 | 5.754 | 0.421 | 5.754 |
| 10 | 0.2 | 1.00 | 0 | 0 | 6.465 | 0.439 | 6.465 |
| 11 | 0.2 | 1.00 | 0 | 0 | 7.327 | 0.421 | 7.327 |
| 12 | 0.2 | 1.00 | 0 | 0 | 8.652 | 0.421 | 8.652 |

Weights in kilograms

## B)

MFYPR version 2a
Run: yr5a
Index file 3/5/2003
Time and date: 19:50 05/05/03
Fbar age range: 4-8

| Age | M |  | Mat | PF |  | PM | SWt | Sel |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3 | 0.2 | 0.11 | 0 | 0 | 1.316 | CWt |  |  |
| 4 | 0.2 | 0.26 | 0 | 0 | 1.822 | 0.172 | 1.316 |  |
| 5 | 0.2 | 0.48 | 0 | 0 | 2.444 | 0.295 | 2.842 |  |
| 6 | 0.2 | 0.74 | 0 | 0 | 3.183 | 0.378 | 3.183 |  |
| 7 | 0.2 | 0.91 | 0 | 0 | 4.040 | 0.397 | 4.040 |  |
| 8 | 0.2 | 0.98 | 0 | 0 | 4.956 | 0.417 | 4.956 |  |
| 9 | 0.2 | 0.99 | 0 | 0 | 5.754 | 0.431 | 5.754 |  |
| 10 | 0.2 | 1.00 | 0 | 0 | 6.465 | 0.449 | 6.465 |  |
| 11 | 0.2 | 1.00 | 0 | 0 | 7.327 | 0.442 | 7.327 |  |
| 12 | 0.2 | 1.00 | 0 | 0 | 8.652 | 0.442 | 8.652 |  |

Weights in kilograms

Table 2.5.6.3 Saithe in the Faroes (Division Vb). Yield-per-recruit. A) summary table based on the run where 1998 year class was reset to approximately the highest observed B) summary table based on the predicted 1998 year-class strength from the XSA.

MFYPR version 2a
Run: yr2
Time and date: 19:46 05/05/03

| Yield per results |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FMult | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | SpwnNosSpwn | SSBSpwn |
| 0.0000 | 0.0000 | 0.0000 | 0.0000 | 5.5167 | 22.5268 | 3.4792 | 18.7235 | 3.4792 | 18.7235 |
| 0.1000 | 0.0330 | 0.1228 | 0.5760 | 4.9050 | 18.1525 | 2.8908 | 14.4189 | 2.8908 |  |
| 0.2000 | 0.0660 | 0.2104 | 0.9164 | 4.4695 | 15.2017 | 2.4774 | 11.5339 | 2.4774 |  |
| 0.3000 | 0.0990 | 0.2763 | 1.1262 | 4.1421 | 13.1010 | 2.1712 | 9.4956 | 2.1712 | 11.5339 |
| 0.4000 | 0.1321 | 0.3279 | 1.2592 | 3.8860 | 11.5434 | 1.9353 | 7.9971 | 1.9353 |  |
| 0.5000 | 0.1651 | 0.3697 | 1.3452 | 3.6793 | 10.3506 | 1.7481 | 6.8603 | 1.7481 | 7.9971 |
| 0.6000 | 0.1981 | 0.4042 | 1.4014 | 3.5086 | 9.4128 | 1.5960 | 5.9758 | 1.5960 | 5.8603 |
| 0.7000 | 0.2311 | 0.4334 | 1.4382 | 3.3646 | 8.6590 | 1.4699 | 5.2727 | 1.4699 | 5.9758 |
| 0.8000 | 0.2641 | 0.4584 | 1.4621 | 3.2412 | 8.0416 | 1.3637 | 4.7036 | 1.3637 | 4.7036 |
| 0.9000 | 0.2971 | 0.4802 | 1.4773 | 3.1341 | 7.5277 | 1.2731 | 4.2358 | 1.2731 | 4.2358 |
| 1.0000 | 0.3301 | 0.4994 | 1.4866 | 3.0399 | 7.0939 | 1.1949 | 3.8460 | 1.1949 | 3.8460 |
| 1.1000 | 0.3632 | 0.5164 | 1.4917 | 2.9564 | 6.7231 | 1.1267 | 3.5172 | 1.1267 | 3.5172 |
| 1.2000 | 0.3962 | 0.5317 | 1.4940 | 2.8815 | 6.4026 | 1.0666 | 3.2370 | 1.0666 | 3.2370 |
| 1.3000 | 0.4292 | 0.5455 | 1.4943 | 2.8141 | 6.1229 | 1.0134 | 2.9959 | 1.0134 | 2.9959 |
| 1.4000 | 0.4622 | 0.5581 | 1.4932 | 2.7528 | 5.8766 | 0.9660 | 2.7867 | 0.9660 | 2.7867 |
| 1.5000 | 0.4952 | 0.5696 | 1.4911 | 2.6968 | 5.6581 | 0.9234 | 2.6037 | 0.9234 | 2.6037 |
| 1.6000 | 0.5282 | 0.5801 | 1.4884 | 2.6454 | 5.4629 | 0.8849 | 2.4426 | 0.8849 | 2.4426 |
| 1.7000 | 0.5612 | 0.5899 | 1.4851 | 2.5980 | 5.2873 | 0.8500 | 2.2999 | 0.8500 | 2.2999 |
| 1.8000 | 0.5943 | 0.5989 | 1.4816 | 2.5541 | 5.1284 | 0.8182 | 2.1726 | 0.8182 | 2.1726 |
| 1.9000 | 0.6273 | 0.6074 | 1.4779 | 2.5132 | 4.9840 | 0.7890 | 2.0586 | 0.7890 | 2.0586 |
| 2.0000 | 0.6603 | 0.6153 | 1.4741 | 2.4751 | 4.8520 | 0.7623 | 1.9560 | 0.7623 | 1.9560 |


| Reference point | F multiplier | Absolute F |
| :--- | :---: | :---: |
| Fbar(4-8) | 1.0000 | 0.3301 |
| FMax | 1.2671 | 0.4183 |
| F0.1 | 0.4792 | 0.1582 |
| F35\%SPR | 0.5321 | 0.1757 |
| $\quad$ Flow | 0.3152 | 0.1041 |
| $\quad$ Fmed | 1.0585 | 0.3494 |
| $\quad$ Fhigh | 2.4647 | 0.8137 |

Weights in kilograms
A)
B)

MFYPR version 2a
Run: yr1a
Time and date: 18:48 03/05/03
Yield per results

| FMult | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | SpwnNosSpwn | SSBSpwn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.0000 | 0.0000 | 0.0000 | 0.0000 | 5.5167 | 22.4316 | 3.4792 | 18.6509 | 3.4792 | 18.6509 |
| 0.1000 | 0.0373 | 0.1371 | 0.6386 | 4.8341 | 17.5654 | 2.8226 | 13.8619 | 2.8226 | 13.8619 |
| 0.2000 | 0.0747 | 0.2305 | 0.9876 | 4.3695 | 14.4558 | 2.3826 | 10.8248 | 2.3826 | 10.8248 |
| 0.3000 | 0.1120 | 0.2988 | 1.1889 | 4.0304 | 12.3248 | 2.0669 | 8.7619 | 2.0669 | 8.7619 |
| 0.4000 | 0.1493 | 0.3513 | 1.3090 | 3.7704 | 10.7879 | 1.8292 | 7.2893 | 1.8292 | 7.2893 |
| 0.5000 | 0.1867 | 0.3931 | 1.3823 | 3.5636 | 9.6353 | 1.6436 | 6.1973 | 1.6436 | 6.1973 |
| 0.6000 | 0.2240 | 0.4274 | 1.4272 | 3.3944 | 8.7433 | 1.4946 | 5.3625 | 1.4946 | 5.3625 |
| 0.7000 | 0.2613 | 0.4561 | 1.4547 | 3.2528 | 8.0349 | 1.3725 | 4.7083 | 1.3725 | 4.7083 |
| 0.8000 | 0.2987 | 0.4806 | 1.4710 | 3.1321 | 7.4600 | 1.2704 | 4.1849 | 1.2704 | 4.1849 |
| 0.9000 | 0.3360 | 0.5019 | 1.4801 | 3.0278 | 6.9849 | 1.1838 | 3.7585 | 1.1838 | 3.7585 |
| 1.0000 | 0.3733 | 0.5205 | 1.4845 | 2.9364 | 6.5859 | 1.1095 | 3.4059 | 1.1095 | 3.4059 |
| 1.1000 | 0.4107 | 0.5370 | 1.4857 | 2.8555 | 6.2462 | 1.0450 | 3.1104 | 1.0450 | 3.1104 |
| 1.2000 | 0.4480 | 0.5518 | 1.4849 | 2.7832 | 5.9534 | 0.9885 | 2.8598 | 0.9885 | 2.8598 |
| 1.3000 | 0.4853 | 0.5652 | 1.4827 | 2.7181 | 5.6985 | 0.9386 | 2.6451 | 0.9386 | 2.6451 |
| 1.4000 | 0.5227 | 0.5773 | 1.4796 | 2.6591 | 5.4745 | 0.8942 | 2.4595 | 0.8942 | 2.4595 |
| 1.5000 | 0.5600 | 0.5884 | 1.4760 | 2.6052 | 5.2759 | 0.8544 | 2.2977 | 0.8544 | 2.2977 |
| 1.6000 | 0.5973 | 0.5986 | 1.4719 | 2.5558 | 5.0985 | 0.8186 | 2.1556 | 0.8186 | 2.1556 |
| 1.7000 | 0.6347 | 0.6080 | 1.4677 | 2.5103 | 4.9391 | 0.7862 | 2.0299 | 0.7862 | 2.0299 |
| 1.8000 | 0.6720 | 0.6167 | 1.4633 | 2.4681 | 4.7949 | 0.7567 | 1.9182 | 0.7567 | 1.9182 |
| 1.9000 | 0.7093 | 0.6249 | 1.4590 | 2.4289 | 4.6638 | 0.7298 | 1.8182 | 0.7298 | 1.8182 |
| 2.0000 | 0.7467 | 0.6324 | 1.4546 | 2.3923 | 4.5439 | 0.7051 | 1.7283 | 0.7051 | 1.7283 |


| Reference point | F multiplier | Absolute $\mathbf{F}$ |
| :--- | :---: | :---: |
| Fbar(4-8) | 1.0000 | 0.3733 |
| FMax | 1.1041 | 0.4122 |
| F0.1 | 0.413 | 0.1542 |
| F35\%SPR | 0.4667 | 0.1742 |
| $\quad$ Flow | 0.2757 | 0.1029 |
| $\quad$ Fmed | 0.935 | 0.3491 |
| $\quad$ Fhigh | 2.1934 | 0.8189 |

[^3]Table 2.5.6.4 Saithe in the Faroes (Division Vb). Prediction with management table. A) based on the run where 1998 year class was reset to approximately the highest observed B) based on the predicted 1998 year-class strength from the XSA.

## A)

MFDP version 1a
Run: man9
Index file 3/5/2003
Time and date: 09:25 07/05/03
Fbar age range: 4-8

2003

| Biomass | SSB | FMult | FBar | Landings |
| :---: | :---: | :---: | :---: | :---: |
| 282175 | 141588 | 1.0000 | 0.4127 | 67423 |


| 2004 <br> Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 249405 | 139504 | 0.0000 | 0.0000 | 0 | 292020 | 185174 |
| . | 139504 | 0.1000 | 0.0413 | 8230 | 282703 | 176854 |
| . | 139504 | 0.2000 | 0.0825 | 16035 | 273880 | 168996 |
| . | 139504 | 0.3000 | 0.1238 | 23441 | 265524 | 161572 |
| . | 139504 | 0.4000 | 0.1651 | 30470 | 257606 | 154556 |
| . | 139504 | 0.5000 | 0.2064 | 37144 | 250102 | 147924 |
| . | 139504 | 0.6000 | 0.2476 | 43484 | 242987 | 141654 |
| . | 139504 | 0.7000 | 0.2889 | 49508 | 236238 | 135723 |
| . | 139504 | 0.8000 | 0.3302 | 55235 | 229835 | 130112 |
| . | 139504 | 0.9000 | 0.3715 | 60681 | 223758 | 124803 |
| . | 139504 | 1.0000 | 0.4127 | 65862 | 217987 | 119777 |
| . | 139504 | 1.1000 | 0.4540 | 70793 | 212506 | 115018 |
| . | 139504 | 1.2000 | 0.4953 | 75489 | 207299 | 110511 |
| . | 139504 | 1.3000 | 0.5366 | 79961 | 202349 | 106240 |
| . | 139504 | 1.4000 | 0.5778 | 84223 | 197642 | 102193 |
| . | 139504 | 1.5000 | 0.6191 | 88285 | 193165 | 98357 |
| . | 139504 | 1.6000 | 0.6604 | 92160 | 188905 | 94719 |
| . | 139504 | 1.7000 | 0.7016 | 95857 | 184849 | 91269 |
| . | 139504 | 1.8000 | 0.7429 | 99385 | 180987 | 87995 |
| . | 139504 | 1.9000 | 0.7842 | 102755 | 177308 | 84888 |
| . | 139504 | 2.0000 | 0.8255 | 105974 | 173801 | 81939 |

Input units are thousands and kg - output in tonnes

Table 2.5.6.4. Continued

## B)

MFDP version 1a
Run: man9a
Index file 3/5/2003
Time and date: 09:33 07/05/03
Fbar age range: 4-8

| 2003 <br> Biomass | SSB | FMult | FBar | Landings |
| :---: | :---: | :---: | :---: | :---: |
| 322257 | 155437 | 1.0000 | 0.4097 | 72844 |


| 2004 <br> Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{2 8 5 0 1 7}$ | 159676 | 0.0000 | 0.0000 | 0 | 331006 | 215199 |
| . | 159676 | 0.1000 | 0.0410 | 9367 | 320333 | 205710 |
| . | 159676 | 0.2000 | 0.0819 | 18263 | 310215 | 196733 |
| . | 159676 | 0.3000 | 0.1229 | 26713 | 300618 | 188239 |
| . | 159676 | 0.4000 | 0.1639 | 34743 | 291513 | 180200 |
| . | 159676 | 0.5000 | 0.2049 | 42377 | 282873 | 172589 |
| . | 159676 | 0.6000 | 0.2458 | 49637 | 274671 | 165381 |
| . | 159676 | 0.7000 | 0.2868 | 56544 | 266882 | 158554 |
| . | 159676 | 0.8000 | 0.3278 | 63116 | 259483 | 152085 |
| . | 159676 | 0.9000 | 0.3688 | 69374 | 252453 | 145954 |
| . | 159676 | 1.0000 | 0.4097 | 75333 | 245770 | 140143 |
| . | 159676 | 1.1000 | 0.4507 | 81010 | 239416 | 134631 |
| . | 159676 | 1.2000 | 0.4917 | 86421 | 233372 | 129404 |
| . | 159676 | 1.3000 | 0.5326 | 91579 | 227622 | 124445 |
| . | 159676 | 1.4000 | 0.5736 | 96500 | 222149 | 119738 |
| . | 159676 | 1.5000 | 0.6146 | 101194 | 216938 | 115271 |
| . | 159676 | 1.6000 | 0.6556 | 105675 | 211975 | 111028 |
| . | 159676 | 1.7000 | 0.6965 | 109953 | 207247 | 106999 |
| . | 159676 | 1.8000 | 0.7375 | 114040 | 202740 | 103171 |
| . | 159676 | 1.9000 | 0.7785 | 117945 | 198444 | 99534 |
| . | 159676 | 2.0000 | 0.8195 | 121679 | 194346 | 96076 |

Input units are thousands and kg - output in tonnes

2.5.1.1 Saithe in the Faroes (Division Vb). Landings in 1000 tonnes.


Figure 2.5.3.1 Saithe in the Faroes (Division Vb ). Mean weight $(\mathrm{kg})$ at age in the catches in 1961-2002.



Figure 2.5.4.1 Saithe in the Faroes (Division Vb). Observed (upper figure) and fitted values (lower figure) proportion mature-at-age for the period 1983-2002.


Figure 2.5.5.1
Saithe in the Faroes (Division Vb). Log-catchability residuals for age groups 3 and 5-11 from XSA.


Figure 2.5.5.2 Saithe in the Faroes (Division Vb). Retrospective analysis of average fishing mortality of age groups 4-8 from XSA for the years 1996-2002.


Figure 2.5.5.3 Saithe in the Faroes (Division Vb). Retrospective analysis of spawning stock biomass of age groups 4-8 from XSA for the years 1996-2002.


Figure 2.5.5.4 Saithe in the Faroes (Division Vb). Fishing mortality (average F ages 4-8).


Figure 2.5.5.5 Saithe in the Faroes (Division Vb ). Recruitment at age 3 (millions).


Figure 2.5.5.6 Saithe in the Faroes (Division Vb). Spawning stock biomass ( 1000 tonnes).


Figure 2.5.5.7 Saithe in the Faroes (Division Vb). Stock-Recruitment plot.

## SPRING SURVEY

SUMMER SURVEY














Figure 2.5.5.8 Saithe in the Faroes (Division Vb). Length distribution from spring (s) and summer survey 19962002. NB! Different scale for year 2001 summer survey.



MFYPR version 2a
Run: yr2
Time and date: 19:46 05/05/03

|  |  |  |
| :--- | :---: | :---: |
| Reference point | F multiplier | Absolute |
| Fbar(4-8) | 1.0000 | 0.3301 |
| FMax | 1.2671 | 0.4183 |
| F0.1 | 0.4792 | 0.1582 |
| F35\%SPR | 0.5321 | 0.1757 |
| Flow | 0.3152 | 0.1041 |
| Fmed | 1.0585 | 0.3494 |
| Fhigh | 2.4647 | 0.8137 |
|  |  |  |

MFDP version 1 a
Run: man9
Time and date: 09:25 07/05/03
Fbar age range: 4-8

Input units are thousands and kg - output in tonnes

Figure 2.5.6.1 Saithe in the Faroes (Division Vb). Fish stock summary.


Figure 2.5.6.2 Saithe in the Faroes (Division Vb). Stock- recruitment.


Figure 2.5.6.3 Saithe in the Faroes (Division Vb). The history of the stock/fishery in relation to the four reference points.


Figure 2.5.6.4 Saithe in the Faroes (Division Vb). Projected composition in number by year classes in the catch in 2002 (left figure) and the composition in SSB in 2003 by year classes (right figure).

### 3.1 Introduction

### 3.1.1 Description of the fisheries

Demersal fisheries take place all around Iceland including variety of gears and boats of all sizes. The most important fleets targeting demersal fish stocks are given below and the spatial distribution of the fisheries are shown in figure 3.1. to 3.3.

- Large and small trawler using demersal trawl. This fleet is the most important one fishing cod, haddock, saithe, redfish as well as a number of other species. This fleet is operating year around, mostly outside 12 nautical miles form the shore.
- Boats (<300 GRT) using gillnets. These boats are mostly targeting cod but cod haddock and a number of other species are included. This fleet is mostly operating close to the shore.
- Boats using longlines. These boats are both small boats ( $<10$ GRT) operating in shallow waters as well as much larger vessels operating in deeper waters. Cod and haddock are the main target species of this fleet but a number of less important species are also caught, some of them in directed fisheries.
- Boats using handlines. These are small boats around 10 GRT and about $300<6$ GRT boats operating in a effort control system where each boat is allocated certain number of days for each year. Cod is the most important target species of this fleet with saithe following as the second most important species.
- Boats using danish seine. (20-300 GRT) The most important species for this fleet are cod and haddock but this fleet is the most important fleet fishing for a variety of flat fishes like plaice, dap and witch.

In addition to those fleets a number of other fleets targeting inverberates and pelagic fishes can affect demersal fish stocks, both through discard other hidden mortality.

### 3.1.2 Regulation of Demersal Fisheries

With the extension of the fisheries jurisdiction to 200 miles in 1975, Iceland introduced new measures to protect young juvenile fish. The mesh size in trawls was increased from 120 mm to 155 mm in 1977. Mesh size of 135 mm was only allowed in the fisheries for redfish in certain areas. Since 1998 a mesh size of 135 is allowed in the codend in all trawl fisheries not using "Polish cover". In 1977 a system was implemented whereby fishing can be forbidden immediately in areas where the number of small fish in the catches exceeds a certain percentage $(25 \%<55 \mathrm{~cm}$ for cod and saithe and $25 \%$ $<48 \mathrm{~cm}$ for haddock). These areas have usually been closed for two weeks and can be extended in time and space by directives if necessary.

A system of transferable boat quotas was introduced in 1984. The agreed quotas were based on the Marine Research Institute's TAC recommendations, taking some socio-economic effects into account. Until 1990, the quota year corresponded to the calendar year but at present the quota, or fishing year, starts on September 1 and ends on August 31 the following year. This was done to meet the needs of the fishing industry.

### 3.1.3 Discards

In recent years discards have received increased attention. Discard has always been a part of the Icelandic fisheries but in recent years, both the magnitude and the age composition of the discards could have been changing, because of changes in economic constraints of the fisheries. Discard is illegal in Icelandic waters except small handliners have been allowed to discard undersized cod.

Since 2000 systematic data collection to estimate discards in demersal fisheries has been carried out. Prior to 2000 discards have been be estimated by comparing length distributions from observers aboard the vessels and from landings. Pálsson (2003) summarises the estimated discard of haddock 1988-2000 and Pálsson et al (2002 and 2003) summarize the discards in Icelandic waters in the main gears targeting cod, haddock, saithe and redfish. The main results of those reports are presented in Tables 3.1 and 3.3.

Tables 3.1 and 3.2 show that estimated discard of cod in 2002 is $1 \%$ of landed catch, decreasing from $1.8 \%$ from last year. Comparison with limited data from earlier years indicates that the discard in 2001 and 2002 is considerably less than in preceding years. The results indicate that the danish seine and gillnet fleet are discarding larger cod than the bottom trawl and longliner fleet.

Estimated haddock discard increased from 2001 to 2002 but is still much less than it was in most years from 1989 to 2000. (Table 3.3). Contradicting trend in Haddock discards in the longline and bottom trawl fishery from 2001 to 2002 is a point of concern as is the high percentage discarded in the danish seine fishery which is also the fleet with highest discard percentage for cod.

Data from 2002 indicates that discard of saithe and redfish was negligible. Discard of golden redfish is a problem in some years but in recent years large areas where small redfish is to be expected have been permanently closed.

A number of fleets that in some years do contribute to discarding are not included in the these studies.

The nephrops fishery takes place off the south coast (Figure 3.1). Small haddock as well as a number of flatfish species are known to have been discarded in this fishery.

The shrimp fishery takes place in many areas around Iceland (Figure 3.3). Small redfish used to be bycatch in the fishery for shrimp off the north coast but since 1995 sorting grid has been mandatory in this fishery. In the last decade no redfish recruitment has on the other hand been seen in those areas. In the shallow water areas off the north coast as well as in the areas off the south west and west coast bycatch of small cod and haddock has often caused a problem in this fishery. For a number of years the inshore shrimp fishery has not been opened if the number of young cod and haddock exceeds certain limit. In the last $3-5$ years the shrimp fishery in most of these areas except the two fjords off the north-west coast has been closed as the shrimp stocks have collapsed, mostly due to cod predation. (anon 2002).

Discards from handliners is not included in the presented studies but handliners are allowed to discard fish below certain size. Recent research (Pálsson et.al 2003a) has though shown that the mortality of cod discarded from handlines is high.

Bycatch of saithe and juvenile cod and haddock is a potential problem in pelagic trawl fisheries. (Figure 3.3). Saithe is by catch in the fishery for blue whiting off the south east coast but small cod and haddock can be a problem in herring fishery in shallower water.

### 3.1.4 Adoption of a Harvest Control Rule for the Icelandic cod stock in 1995

In May 1995, the Icelandic government adopted a Harvest Control Rule (HCR) for the Icelandic cod fishery, based on work carried out by a government appointed group of fisheries scientists and economists (Anon., 1994; Baldursson et al., 1996; Daníelsson et al., 1997). The group investigated the consequences of various long-term harvesting stategies for cod by using risk analysis, taking into account biological and economic interactions between cod and its major prey, capelin and shrimp. The group showed that a harvest rate of $25 \%$ of the average fishable ( $4+$ ) biomass of cod at the start and the end of assessment year with a minimum of 155 thousand tonnes TAC would lead to a low probability of stock collapse, defined as SSB going below 100 thousand tonnes. The government implemented this catch-rule as a Harvest Control Rule in the next five fishing years.

### 3.1.5 Amendments adopted in June 2000

The assessment of the Icelandic cod stock in the year 2000 showed that the fishable biomass in 2000 had been overestimated by 180 thousand tonnes in the preceding assessment. Based on the 2000 assessment the HCR for the quota year 2000/2001 resulted in a recommended catch of 203 thousand tonnes. This reduction in catch between two consecutive years, which was largely driven by the downward revision in stock estimates, highlighted to the managers the uncertainty in stock assessments and the undesirability of tying a catch rule directly to point estimators in stock assessment. In June 2000 the Icelandic government therefore asked the MRI to explore whether an upper limit of between-year changes in TAC (catch-stabilizer) would jeopardise the original aim of the long-term harvesting strategy imposed by the HCR, with the addition of excluding the 155 thousand tonnes TAC floor.

Under the given time constraint only limited studies were possible. The basic approach taken was the same as that done previously by the working group (Stefánsson et al. 1997a; 1997b) and the work was carried out by one of its member. In addition to simulating cod, capelin and shrimp the analysis included two seal species and three species of baleen whales. The same criterion was used for the definition of stock collapse i.e SSB going below 100 thousand tonnes. No
density dependent growth in the cod stock was assumed and only limited options of catch developments of whales and seals were explored, but different assumptions will affect the mean catch figures of cod. Fifteen percent CV in stock estimates was assumed. The general conclusion of all base-case trials showed limited sensitivity of introduction of a range of catch-stabilizers (10-60 thous tons). However, when various catch-stabilisers were applied under a regime of drastic reduction in recruitment (half the normal recruitment per SSB ), the effects became clear; the lower the stabiliser was fixed, the greater probability of SSB collapse. It appeared that when catch-stabiliser applied was 25 thous tonnes or less, the risk increased significantly, while catch-stabiliser, allowing 30 thous tonnes or higher interannual changes in catches performed far better. In light of these provisional trials, the 30 thous tonnes catch-stabiliser was considered a safe approach.

On the basis of these results the Icelandic government adopted a modification to the HCR by including a 30 thousand tonnes catch-stabiliser and abandoning the minimum catch floor of 155 thousand tonnes. This resulted in a TAC of 220 thousand tonnes for the fishing year 2000/2001 instead of 203 thousand tonnes and 190 thousand tonnes for the fishing year 2001/2002 instead of 155 thousand tonnes if no stabiliser would have been in effect.

At the time of the catch-rule amendment, because of time constrains, detailed alternative simulations were not possible. A working group was set up by the Ministry of Fisheries in 2001 with the objectives to analyse the experience of using the catch rule and try out alternative approaches taking into account obvious shortcomings of the current harvest control rule and use state of the art knowledge for further development. This working group was supposed to deliver a preliminary report before the start of the fishing year 2002/2003 but at present no report has been presented.

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Landings, discards in numbers (thousand fishes) and weight (tonnes) and as a proportion (\%) of landings, by species and gear for the year 2001. From Pálsson et.al (2002).

| Species | Gear | Discards <br> (Thous. Fishes) | Discard <br> (tons) | DL $_{50}$ <br> cm | Discard- <br> Percentage weight |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Cod | Longlines | 600 | 464 | 43.7 | 1.0 |
| Cod | Gillnets | 560 | 1620 | 53.9 | 3.0 |
| Cod | Danish seine | 903 | 1259 | 53.7 | 7.6 |
| Cod | Bottom trawl | 649 | 471 | 42.8 | 0.5 |
| Cod | Total | $\mathbf{2 7 1 2}$ | $\mathbf{3 8 1 4}$ |  | $\mathbf{1 . 8}$ |
| Haddock | Longlines | 1391 | 560 | 35.7 | 4.6 |
| Haddock | Bottom trawl | 989 | 456 | 37.9 | 2.1 |
| Haddock | Total | $\mathbf{2 3 8 0}$ | $\mathbf{1 0 1 6}$ |  | $\mathbf{3 . 0}$ |

Table 3.2
Landings, discards in numbers (thousand fishes) and weight (tonnes) and as a proportion (\%) of landings, by species and gear for the year 2002. From Pálsson et.al (2003)

| Species | Gear | Catch <br> (tonn) | Discards <br> (Thous. <br> Fishes) | Discard <br> (tons) | DL $_{50}$ <br> cm | Discard- <br> Percentage |
| :---: | :--- | :---: | :---: | :---: | :---: | :---: |
| Cod | Longlines | 42154 | 220 | 124 | 40.4 | 0.3 |
| Cod | Gillnets | 44162 | 220 | 515 | 48.3 | 1.2 |
| Cod | Danish seine | 13575 | 694 | 987 | 49.6 | 7.3 |
| Cod | Bottom trawl | 85740 | 323 | 196 | 41.7 | 0.2 |
| Cod | Total | $\mathbf{1 8 5 6 3 1}$ | $\mathbf{1 4 5 7}$ | $\mathbf{1 8 2 2}$ |  | $\mathbf{1 . 0}$ |
| Haddock | Longline | 13568 | 311 | 151 | 37.2 | 1.1 |
| Haddock | Danish seine | 3582 | 466 | 382 | 40.5 | 10.7 |
| Haddock | Bottom trawl | 29883 | 2908 | 1782 | 41.1 | 6.0 |
| Haddock | Total | $\mathbf{4 7 0 3 3}$ | $\mathbf{3 6 8 5}$ | $\mathbf{2 3 1 5}$ |  | $\mathbf{4 . 9}$ |
| Saithe | Bottom trawl | 35260 | + | + |  | + |
| Redfish | Bottom trawl | 51633 | 0 | 0 |  | 0 |

Table 3.3 Haddock discards in the Icelandic demersal trawl fishery 1988-2000 From Pálsson (2002).

| Year | Discards by numbers |  | Discards by weight |  | $\mathrm{DL}_{50}$(cm) | Yearclass $^{5}$ | Stocksize ${ }^{5}$$(\text { age } 3+)^{4}$ | Landings ${ }^{5}$ <br> Numbers ${ }^{3}$ | Landings <br> Tons |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | (millions) | (\%) ${ }^{\text {T }}$ | (tons) | $(\%)^{2}$ |  |  |  |  |  |
| 1988 | 2.9 | 11.7 | 1481 | 3.8 | 40.1 | 47 | 151 | 24.8 | 39088 |
| 1989 | 2.6 | 9.9 | 1499 | 3.4 | 41.0 | 24 | 168 | 26.3 | 44215 |
| 1990 | 1.1 | 3.5 | 364 | 0.8 | 37.1 | 22 | 145 | 31 | 47158 |
| 1991 | 6.9 | 37.7 | 3349 | 9.7 | 43.8 | 79 | 120 | 18.3 | 34661 |
| 1992 | 8.4 | 47.5 | 3858 | 13.3 | 39.3 | 169 | 106 | 17.7 | 29093 |
| 1993 | 5.7 | 25.1 | 2414 | 8.0 | 38.5 | 37 | 129 | 22.7 | 30132 |
| 1994 | 8.9 | 29.6 | 4236 | 10.7 | 40.4 | 41 | 127 | 30.1 | 39474 |
| 1995 | 12.8 | 44.1 | 8397 | 19.6 | 42.2 | 75 | 119 | 29 | 42829 |
| 1996 | 10.2 | 39.4 | 4577 | 11.6 | 41.1 | 38 | 108 | 25.9 | 39466 |
| 1997 | 8.3 | 50.6 | 6160 | 22.3 | 44.4 | 89 | 89 | 16.4 | 27643 |
| 1998 | 5.4 | 34.6 | 2501 | 10.3 | 40.3 | 18 | 96 | 15.6 | 24191 |
| 1999 | 2.6 | 14.1 | 1349 | 5.2 | 38.8 | 80 | 91 | 18.4 | 25960 |
| 2000 | 3.4 | 21.8 | 1930 | 8.4 | 40.5 | 80 | 86 | 15.6 | 22990 |
| 2001 | 1.0 |  | 456 | 2.1 |  |  |  |  |  |
| 2002 | 2.9 |  | 1782 | 6.0 |  |  |  |  | 29883 |

With reference to numbers ${ }^{1}$ or weights ${ }^{2}$ landed. ${ }^{3}$ Millions. ${ }^{4}$ Thousand tons. ${ }^{5}$ Source: Anon. (2000) The 2001 and 2002 values are obtained from Pálsson et.al (2002 and 2003).

Danısh seine ettort settıngs/square mile/year


Figure 3.1 Distribution of effort in the Icelandic Danish seine and Nephrops fishery 1993-2002.

Gillnet eftort nets/square mile/year


Longline effort 1000 hooks/square mile /year


Handline effort days/square mile/year


Figure 3.2
Distribution of effort in the Icelandic gillnet fishery 1993 - 2002, longline fishery 2000-2002 and handline fishery 2000-2002

Bottom trawl ettort hours/square mile/year


Figure 3.3
Distribution of effort in the Icelandic bottom trawl fishery 1993-2002, shrimp fishery 1993-2002 and pelagic trawl fishery 2000-2002.

This stock was not assessed by the working group this year

### 3.3 Icelandic cod (Division Va)

### 3.3.1 Stock definition

The Icelandic cod stock is distributed all around Iceland and in the assessment it is assumed to be a single homogenous unit. Main spawning takes place in late winter mainly off the southwest coast but smaller regional spawning components have also been observed off the west, north, and east coasts. The pelagic eggs and larvae from the main spawning grounds drift clockwise around the island to the main nursery grounds off the north coast. A larval drift to Greenland waters has been recorded in some years and substantial immigrations of mature cod from Greenland have been observed in some years which are assumed to be of Icelandic origin. Such migration was last observed in 1990 from the 1984 year class, about 30 millions 6 years old in 1990. Extensive tagging in the last century and during recent years shows no indication of significant emigration from Iceland to other areas.

### 3.3.2 Fishery

The fleet fishing for cod at Iceland operates throughout the year. The fishing vessels can be grouped into three main categories: 1) Multi-gear boats; < 300 GRT, 2) Small boats; < 20 GRT, 3) Trawlers; > 300 GRT.

The trawlers operate throughout the year outside the 12 mile limits. They follow spawning and feeding migration patterns of cod and fish on spawning grounds off the south west and south-coasts during the spawning season but move to the feeding areas off the northwest coast during the summer time. During the autumn, this fleet is more spread out. The multi-gear boats operate mainly using gillnet during the spawning season in winter and spring along the south-west coasts but in recent years this fleet has also used gillnet in late autumn. In the years 1995 to 1998 this fleet increased the mesh size in their nets from 7 to 9 inches but reduced the mesh size back to 8 inches in 1999. During the last 4 years the bulk of the gillnet catches are taken in 8 and 9 inches mesh size (Figure 3.3.3). Part of this fleet uses longlines during autumn and early winter. During summer some of these boats trawl along the coast out to the 3 mile limit. Others fish with Danish seines close to the shore. Most of the smaller boats operate with handlines, mainly in shallow waters during the summer and autumn period. Landings by gear since 1982 are shown in figure 3.3.2.

### 3.3.3 Data

### 3.3.3.1 Fishery dependent data

### 3.3.3.1.1 Landings

In the period 1978-1981 landings of cod increased from 320000 t to 469000 t due to immigration of the strong 1973 year class from Greenland waters combined with an increase in fishing effort. Catches declined rapidly to only 280000 t in 1983. Although cod catches have been regulated by quotas since 1984, catches increased to 392000 t in 1987 due to the recruitment of the 1983 and 1984 year classes to the fishable stock in those years (Table 3.3.1 and Fig. 3.3.1).

During the period 1988-1996 all year classes entering the fishable stock were well below average, or even poor, resulting in a continuous decline in the landings. The 1995 catch of only 170000 t is the lowest since 1942. From 1995 catches increased continuously to 1999 when the estimated landings were 260000 tonnes but decreased to 235000 tonnes in the years 2000 and 2001 and the recorded landings in 2002 were 209000 tonnes.

### 3.3.3.1.2 Sampling intensity

The data samples comprising the age-length keys for 2002 are given in the following table:

| Gear | Season | Area | No. length samples | No. length measured | No.age samples | no.aged aged | No. wt samples |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Bottom trawl | Jan-May | South | 138 | 16270 | 16 | 770 | 526 |
| Bottom trawl | Jan-May | North | 46 | 6666 | 8 | 449 | 298 |
| Danish seine | Jan-May | South | 211 | 35594 | 2570 | 3916 | 3667 |
| Danish seine | Jan-May | North | 29 | 4282 | 1005 | 1487 | 1437 |
| Handlines | Jan-May | South | 0 | 0 | 0 | 0 | 0 |
| Handlines | Jan-May | North | 1 | 150 | 1 | 50 | 50 |
| Gillnet | Jan-May | South | 106 | 7521 | 6 | 299 | 299 |
| Gillnet | Jan-May | North | 14 | 524 | 0 | 0 | 0 |
| Longline | Jan-May | South | 487 | 24048 | 28 | 1369 | 690 |
| Longline | Jan-May | North | 620 | 69459 | 41 | 2180 | 698 |
| Bottom trawl | June-Dec | South | 99 | 5944 | 14 | 665 | 422 |
| Bottom trawl | June-Dec | North | 150 | 17562 | 16 | 795 | 446 |
| Danish seine | June-Dec | South | 48 | 4964 | 5 | 264 | 264 |
| Danish seine | June-Dec | North | 2 | 435 | 0 | 0 | 0 |
| Handlines | June-Dec | South | 9 | 1320 | 3 | 147 | 50 |
| Handlines | June-Dec | North | 39 | 8096 | 6 | 276 | 228 |
| Gillnet | June-Dec | South | 138 | 5776 | 6 | 296 | 296 |
| Gillnet | June-Dec | North | 63 | 4762 | 1 | 50 | 50 |
| Longline | June-Dec | South | 312 | 4198 | 16 | 777 | 532 |
| Longline | June-Dec | North | 641 | 62958 | 55 | 2759 | 1034 |
| Total |  |  | 3153 | 280529 | 3797 | 16549 | 10987 |

In recent years emphasis has been put on relating the sampling scheme to the landings database automatically, calling for samples when certain amount has been landed in each cell, calculated daily ("real time proportional sampling scheme").

### 3.3.3.1.3 Catch in numbers-at-age

Catch in number-at-age is calculated by splitting the landings by 5 fleets, 2 areas and 2 seasons. The gears are long lines, bottom trawl, gillnets, hand lines and Danish seine, seasons January-May (spawning season) and June-December and regions North and South. Historically, there have been some changes in fleet definitions and thus there does not currently exist a fully consistent set of catch-at-age data on a per-fleet basis. In some cases samples are not available for a cell or are to few to give reliable keys. In those cases otolith samples from "related" cells are used. Notably hand lines are included with long lines in the same area and season.

The total catch-at-age data is given in Table 3.3.3 and Figure 3.3.4. The Shephard Nicholson model gives a CV of 0.2 for age groups 4-10. It should be noted that much higher proportions of the older age groups are taken during the first part of the year and this fishing mortality affects estimation of the spawning stock at spawning time. Since the catch-at-age data have historically only been available for January to May, and not by shorter periods, it is assumed that $60 \%$ of those catches were taken during January to March, i.e., before spawning time (Table 3.3.4). Natural mortality before spawning is assumed to be one fourth of the annual natural mortality.

### 3.3.3.1.4 Mean weight-at-age in the landings

Mean weight-at-age in the landings is calculated with the catch in numbers. Before 1993 weighting of cod was relatively uncommon so length-weight relationships were based on little data. Since 1994 weighting has been much more extensive but currently all fishes sampled for otolith are weighted and length-weight relationships can be calculated from current data. The mean weights-at-age in the landings are shown in table 3.3.5 and figure 3.3.7.

Mean weight-at-age have been shown to correlate well with the size of the capelin stock and capelin stock size has been used as a predictor of weights in the landings since 1991. In 1981-1982 weights were low following collapse of the capelin stock and were also relatively low in 1990-1991 when the capelin stock was small. The weights were high in 1994 to 1998 but have been around long-term average since 1999. The observed mean weights-at-age in 2002 were about same as in 2001.

Mean weights-at-age are not available on an annual basis for catches taken before 1973, and hence the average for the years 1973-1991 is used as the constant (in time) mean weight-at-age for earlier years.

Weights-at-age in the landings have been used without modification to compute stock biomasses, with the exception of the spawning stock biomass (see below).

### 3.3.3.1.5 Mean weight-at-age in the landings at spawning time

Weight-at-age data from the commercial catch period January-May have been used for estimation of mean weights-at-age in the spawning stock. It is assumed that catches in different gears and areas appropriately reflect stock composition with regard to mean weight-at-age. Weights in the SSB decreased in 1999 and 2000 after being very high in 1996 to 1998. In 2001 and 2002 an increase was observed for all relevant age groups and the mean weights in 2002 are around long-term average (1974-2001). The peak in 1996 to 1998 could be related to selection of the commercial fleets who were using large mesh size in gillnets in this period (Figure 3.3.3). Mean weights in the spawning stock are shown in table 3.3.6 and figure 3.3.8.

### 3.3.3.1.6 Maturity-at-age at spawning time

Maturity-at-age is based on samples from the commercial fleets in January-May (ICES 1992/Assess:14) (Table 3.3 .7 and Figure 3.3.6). It has been pointed out that using data collected throughout the year may bias the proportion mature in various ways (Stefánsson, 1992). The approach taken is, therefore, to compute the proportion mature at the time of spawning, by considering only the first part of the year (January-May), but aggregating across gears and regions. Maturity-at-age increased substantially from 1982-1995 to relatively high values and decreased again in 1996-2000 but a sharp increase was observed for age groups 3-7 in 2001 and about the same values were observed in 2002.

Maturity-at-age data are not available on an annual basis for catches taken prior to 1973 and, hence, the average for the years 1973-1991 is used as a constant (in time) maturity-at-age for the years prior to 1973.

### 3.3.3.1.7 CPUE

Logbooks were kept on voluntary basis until 1991 and only part of the fleet, mainly trawlers, did send in logbooks. After 1991 logbooks are available from all vessel and gears except for boats less than 10 GRT which kept logbooks on voluntary basis until 1999 but since then also mandatory. Substantial linear trend in catchability in cpue from commercial fleets has been observed (WD-31, NWWG 2002) and they are therefore not used for calibration of assessment models.

The unstandardised CPUE indices and effort from the commercial fleets since 1991 is presented in Figures 3.3.9. A and Tables 3.3.2. In the years 1993-1995 a marked reduction in effort and increase in CPUE was observed with the adoption of the HCR. The largest reduction was by the trawlers who diverted their effort to other species and other areas. The effort increased and CPE decreased in all gears in 1998-2001. In 2002 a decrease in effort and increase in CPUE was observed for all gears expect for gillnets where small decrease in 2002 is observed in CPUE. The increase in effort in 1998-2001 can be explained by overestimation of the stock and the amendment of the HCR in the year 2000.

### 3.3.3.2 Fishery independent data

### 3.3.3.2.1 Survey abundance indices

Since 1985 the Icelandic groundfish survey (IGS) has been carried out annually in March, covering the continental shelf waters around Iceland with 540-600 "semi randomly" distributed fixed stations (Pálsson et al, 1989). The survey design was based on historical information about spatial distribution of cod. Each year 4-5 similar commercial trawlers have been hired to cover the stations using standardised 105 -feet bottom trawl. The horizontal net opening is estimated to be about 17 m and vertical opening about 2.5 m . The standard towing distance is 4 nautical miles.

A conventional Cochran type method was used for calculating survey indices. The strata used follow depth contours. The Cochran indices were calculated separately for two areas: Northern and Southern area and combined. For all models used except for the TSA the indices were combined by simple summation (Table 3.3.8 and figure 3.3.11) but for the TSA tuning the two area indices a weighted geometric mean was calculated (Table3.3.9). The total biomass index from the survey is presented in figure 3.3.10. The Shephard Nicholson model gives a CV of 0.24 for age groups 2-9 for the survey indices.

Figure 3.3.13 show plots of survey index for cod vs. the index of the same year class in the survey one year later. This type of plot should show good relationship if the survey is consistent, except when fishing effort varies much. The best relationship is between ages 3 and 4, age groups that are fully recruited to the survey but age 3 does usually have low fishing mortality.

### 3.3.3.2.2 Mean weight and maturity-at-age in survey

The calculated annual mean weight-at-age in the IGS show similar pattern as the weights in landings although survey weights for age 3 to 5 are always considerably lower than weights from the catches from in the same period. The same applies to the maturity-at-age were much lower values are observed for the younger ages in the survey

Data collected in the Icelandic groundfish survey (IGS) have the potential for providing better estimates of mean weights-at-age in the stock. As the survey takes place early in the year with small meshes in the trawl codend mean weights in the survey multiplied by number in stock would give a much better measure of "real stock size". A problem with using survey weights for calculation of stock biomass is that they are only available back to 1985 and weighting of mean weight-at-age from areas with different growth rates is sensitive to catchability and annual differences in spatial distribution.

### 3.3.4 Stock Assessment

### 3.3.4.1 Recent assessment and reviews

The 2000 assessment showed that the stock had been seriously overestimated in recent years. In May-June 2000 the MRI asked a group of external experts to review and reanalyse the assessment. The group was chaired by Prof. John Pope. Various alternative assessment models were used: XSA (John Pope), Coleraine (Árni Magnússon/Ray Hilborn), Cagean like model (Pat Sullivan), TSA (Guðmundur Guðmundsson) as well as some unconventional methods such as Bormicon (Höskuldur Björnsson). The group met twice, first in May/June 2000 and again in late autumn.

At the spring meeting, several different tuning sets and assumptions were explored. The main conclusion was that the ICES assessment (XSA using a number of cpue tuning series from the survey and the commercial fleets) gave a higher estimate of biomass than most of the assessments done by the reviewing body. Variations in catchability and selection patterns in recent years were identified. The reviewing body did however not suggest in their June meeting that the assessment should be rejected and concluded that the MRI/Working group methodology and procedures were sound.

In the autumn 2000 the reviewing group met in Reykjavík to discuss the results of the review with Icelandic scientists. Various aspects of assessments, data sets and assumptions were discussed. The main critique on the ICES assessment was the use of multiple fleets and area split survey indices in XSA, letting XSA select the weights of different fleets. For future assessment it was suggested that a combined survey index should be used and that XSA should be run on only one tuning series at a time. It was also pointed out that XSA is unable to compensate for changes in efficiency of the commercial fleets (except for down weighting) and there is potential danger of overweighing the commercial cpue indices as they are often highly correlated with catch-at-age data. The reviewing group also recommended the usage of various alternative assessment models.

The 2001 assessment showed that there had been around $15-25 \%$ overestimation in biomass in the years 1998-2000. In 2001 the results from XSA using one survey fleet for calibration was adopted as a final run by ICES. Various other assessment models tried at MRI gave all very similar results. In spring 2001 the Minister of Fisheries asked a group of external experts lead by Dr. Andrew Rosenberg to review the recent years assessments with emphasis on uncertainty in assessments. The group delivered a final report in July 2002
(http://brunnur.stjr.is/interpro/sjavarutv/sjavarutv.nsf/pages/a_rosenberg.html).
The main question posed to the group was "Could this situation have been forseen and prevented?". The main results of the Rosenberg group that assessment of Icelandic cod in the years under revision were driven by high survey indices confirming the conclusion reached by Pope and "it does not seem as if there was any obvious early warning signal of the problems that arose in 1999 and 2000". Simulation studies conducted using XSA with and without shrinkage indicated that a systematic bias was introduced by the usage of shrinkage showing ubiquitous retrospective patterns. Removal of shrinkage, hence bias, improved the retrospective pattern and lowered the precision. The group recommended to use a variety of different assessment methods and test the sensitivity of parameters settings, the use simulation studies and to review the results using retrospective analysis. The group also concluded that the overestimation in recent years should be taken into account in future management consideration.

### 3.3.4.2 Current assessment

Consistent with the above and the results of a study done by Gudmundsson and Jónsson (WD-31, NWWG-2002), showing substantial linear trend in catchability in cpue form commercial fleets, only survey indices were used for calibration of assessment models in the 2002 assessment and six different assessment models were applied. The same approach was followed in the current assessment and five different models were used: XSA and TSA as last year, AD-

CAM- AD-Model builder statistical Catch-at-age Model written and developed at the MRI (WD-33, NWWG-2002), EX-CAM-Statistical Catch-at-age Model written in Excel developed at the MRI (Working Document no. 33), Coleraine-a general statistical catch-at-age model developed at the University of Washington (Working Document no. 31). The last three methods are essentially implementations of closely related models.

### 3.3.4.3 Estimates of fishing mortality

The five different assessment models were run all using the same datasets, catch in number-at-age, Table 3.3.3, and survey indices, Table 3.3.8, expect for TSA using weighted geometric mean of North and South areas indices, Table 3.3.9.

## XSA tuning

Two runs were made by XSA using the same settings. Firstly a run using age groups 3-9 from survey for tuning as in last years assessment and secondly a run using age groups 1-9. To use the latest information available for tuning, the 2003 survey indices were moved three months back in time i.e. to end of December 2002. The resulting tuning diagnostic and terminal F's are presented in Table 3.3.10a and Table 3.3.10b respectively, resulting retrospective analysis in Figure 3.3.16 and Figure 3.3.17 and the log catchability residuals in Figure 3.3.8. The estimated terminal reference F (average of age groups 5-10) is $\mathbf{0 . 7 2}$.

## TSA

The results of the TSA run are presented in Table. 3.3.11. The test statistics from standardised residuals of prediction errors of catches and survey indices seem satisfactory. (Table 3.3.11 and Figure 3.3.18). The results from corresponding retrospective analysis are presented in Figures 3.3.16-17. The terminal reference fishing mortality based on this run is $\mathbf{0 . 6 6}$.

## AD-CAM

The input parameters settings, estimated fishing mortality rates and stock in numbers are presented in Table 3.3.12 along with the resulting residuals. The residuals plot are presented in Figure 3.3.18 a the corresponding retrospective pattern in Figures 3.3.16-17. The terminal reference fishing mortality is estimated $\mathbf{0 . 7 6}$.

## EX-CAM

The estimated parameters and results of the EX-CAM run are presented in Table. 3.3.13 as well as the residuals of prediction errors of catches and survey indices seems satisfactory. (See also Figure 3.3.18 for plot of the residuals). The results from corresponding retrospective analysis are presented in Figures 3.3.16-17. The terminal reference fishing mortality based on this run is $\mathbf{0 . 5 8}$.

## Coleraine

The estimated parameters and results of the Coleraine run are presented in Table. 3.3.14 as well as the residuals of prediction errors of catches and survey indices. (See also Figure 3.3 .18 for plot of the residuals). The results from corresponding retrospective analysis are presented in Figures 3.3.16-17. The terminal reference fishing mortality based on this run is $\mathbf{0 . 4 0}$.

### 3.3.4.4 The selection of a final run

In Table 3.3.15 and Figures 3.3.19 and Figure 3.3.20 a summary of the resulting terminal fishing mortalities and estimated, biomass and stock in numbers in 2003 from the five different models are presented. The estimated stock in weight ( $4+$ ) in the beginning of 2003 from the XSA, TSA, AD-CAM and EX-CAM are similar or in the range of 766-795 thous. tonnes . Those four models also show similar fishing mortality pattern but TSA and EX-CAM estimate somewhat lower F values for the older age groups. The difference in the terminal reference fishing mortalities, $0.58-0.76$, is reflecting the difference in the older ages. Coleraine gives the lowest value for the reference fishing mortalities, 0.4 but there seem to be model configuration problem in this model which could not be resolved. This is reflected in the differences in stock size estimates back in time compared to the results of the other models used (Figure 3.3.21, Working Document no 31).

Comparison of the retrospective results from the various models (Figure 3.3.16-17) show that the most consistent patterns are observed using the AD-CAM model looking at both the reference fishing mortalities and the fishable
biomass $(4+)$. The retrospective pattern from the TSA runs does show the second best consistency and the other models do show somewhat more inconsistent pattern.

In last year assessment the resulting F values from the TSA were used for traditional VPA backwards calculations for the final estimate of the stock in number in the beginning of the assessment year using the RCT3 program to estimate the youngest age groups. This procedure is in many cases unsuitable as the estimates of different age groups are correlated in many models.

In this year the results of the TSA and XSA, the two models used here who have been formally accepted by ICES, give very similar results as the AD-CAM model. The estimated stock in numbers in the beginning of 2003 from TSA and XSA are well within one standard error of the AD-CAM results (Figure 3.3.24).

The NWWG concluded that the AD-CAM modelling approach is the most appropriate since it provides stock and recruitment estimates within the same statistical framework including probability profiles. Medium-term projection is also a natural extension of this type of model approach. Furthermore the AD-CAM model can handle migrations and survey indices in the assessment year and is designed and run by a member of the working group. For these reasons, and for convenience, the AD-CAM run was adopted as a point estimate for forward projections.

The estimated Biomass (4+) in 2003 from the AD-CAM model is 766 thous. tonnes with standard error of 40 . The resulting fishing mortalities are given in Table 3.3.16 and in Figure 3.3.22B. The fishing mortality increased to a peak in 1988, decreased in 1989 but then rose to another peak in 1993. Due to restriction of the cod quota fishing mortality dropped markedly in 1995 and 1996 but has increased since then to $0.76-0.78$ in 2000-2002.

### 3.3.4.5 Stock and recruitment estimates

The resulting stock size in numbers and stock in weight from the final run are given in Tables 3.3.17 and 3.3.19. In the stock in numbers table. The recruitment in the most recent years is estimated by the AD-CAM model. Parameters setting and assumptions made are described in Table 3.3.12.

### 3.3.5 Biological and technical interactions

Several important biological interactions in the ecosystem around Iceland are connected to the cod stock. The single most important interaction is the cod-capelin connection (Pálsson, 1981) and this has been studied in some detail (Magnússon and Pálsson, 1989 and 1991a and Steinarsson and Stefánsson, 1991). Another important interaction is between cod and shrimp. This has been studied by Magnússon and Pálsson (1991b) and Stefánsson et al. (1994). The cod-capelin interaction is used in the short-term prediction in Section 3.3.7.1 based on the results in Steinarsson and Stefánsson (1996).

Various factors affect the natural mortality of cod and several of these factors could change in magnitude in the future. The cod is a cannibal and the mortality through cannibalism has been estimated in Björnsson (WD 26, 1998). Cannibalism occurs mainly on pre-recruits and immature fish. Further, the minke whale, the harbour seal and the grey seal are apex predators, all of which consume cod to varying degrees. Most of these $M$ values will affect cod at an early age, before recruitment to the fishery.

It has been illustrated that not only may cetaceans have a considerable impact on future yields from cod in Division Va (Stefánsson et al., 1995), but seals may have an even greater impact (Stefánsson et al., 1997). These results imply that predictions which do not take into account the possible effects of marine mammals may be too optimistic in terms of long-term yields. It is therefore desirable to include marine mammals as a part of future natural mortality for the cod stock.

A number of fleets operate in Division Va. The primary gears are described in Section 3.3.2. Earlier work by this group included the separation of catches into finer seasonal and areal splits, but this has not been taken further at this meeting.

A numerical description of interactions between fisheries and species requires data on landings as well as catches in numbers-at-age of each species by gear type, region and season.

### 3.3.6.1 Input data to the short-term prediction

For short-term predictions, it is essential to take into account potential changes in mean weights-at-age due to environmental conditions. It has been shown that cod growth is to some extent correlated to size the of the capelin stock. Table 3.3.20 gives the size of the capelin stock biomass since 1979. Regressions based on the capelin stock size are used to predict the mean weights-at-age for age groups 4-8 in the catches and ages 5-8 in the spawning stock for the year 2003. For the year 2004 onwards, the average capelin stock size over the years 1979-2003 is used for prediction. (Table 3.3.24). In the most recent period maturity-at-age has been decreasing but a marked increase was observed in 2001 and 2002. For the short-term predictions the average for the years 2000-2002 has been used for the years 20032005. The exploitation pattern used for the short-term predictions was taken as the average of the years 2000-2002.

Based on the reported landings for the first month of the 2002/2003 fishing year and an assumption of the use of amended harvest control rule for the coming fishing year the expected catch in 2003 will be $210,000 \mathrm{t}$ corresponding to $\mathrm{F}=0.58$.

A TAC constraint is used for this stock since the TAC forecasts have historically been relatively good. The use of last three years average exploitation pattern and a status quo F in 2003 compared to 2002 results in estimated reference fishing mortality of 0.76 and corresponding catches of about 260 thous. tonnes in 2003. This procedure will certainly overestimate the landings in 2003 and most likely also overestimate the F . That would also have been the case if F constraints would have been applied in the two previous assessments by the working group. Further investigation back in time where not done at the meeting. A detailed analysis that takes into account the accuracy and bias of the present estimate is needed to resolve the issue whether a TAC or an F constraint is more appropriate to carry the stock size into the advisory year.

The results from the AD-CAM model were used for recruitment prediction. The RCT3 program was also run with same settings as last year. The combined Cochran survey indices, age groups 1-4 and recruitment estimates from the ADCAM, for the year classes 1981-1998, were used as input for the RCT3 recruitment prediction. The input is given in Table 3.3.22. and the output in Table 3.3.23. The size of the year classes 1998-2002 as estimated by the various models give all very similar estimates, see Table 3.3.15.

### 3.3.6.2 Short-term prediction results

Input data to the short-term prediction and results from projections up to the year 2005 with different management options are presented in Table 3.3.24 and Figure 3.3.23A.

If the buffer of the amended catch control rule (with an upper limit of between year changes in TACs of 30 thous. tonnes) will be applied the resulting TAC in the 2003/2004 fishing year will be 209,000 tonnes. The SSB will increase to about 440 thous. tonnes in 2004 and the resulting reference fishing mortality are about 0.49 . The estimated age distribution of the catches and SSB are shown in figure 3.3.23B

### 3.3.6.3 Input data to the long-term prediction

For long-term predictions, fluctuating environmental conditions can be ignored, but it is essential to take into account potential changes due to density-dependent growth. These have been investigated for this stock (Steinarsson and Stefánsson, 1991 and ICES 1991/Assess:7) where no signs of density-dependent growth were found. However, the results in Schopka (1994) contain indications of some density dependence of growth and this will affect the long-term results at low fishing mortalities. This is not taken into account in typical yield-per-recruit calculations. Effect of catch on mean weight-at-age by selection of the largest individuals of incoming year classes is also an important effect not taken into account.

Naturally, any stock-recruitment relationship will affect yield-potential calculations and this is not taken into account in the yield-per-recruit calculations.

Average exploitation pattern, mean weight-at-age and maturity-at-age over the years 1982-2002 has been used as input (Table 3.3.25).

### 3.3.6.4 Long-term prediction results and biological reference points

The biological reference values for $\mathbf{F}_{\max }$ and $\mathbf{F}_{0.1}$ are 0.38 and 0.20 respectively. Yield-per-recruit at the $\mathbf{F}_{\max }-$ is 1.76 kg . (Figure 3.3.25 Table 3.3.26).

A plot of the spawning stock biomass and recruitment is given in Figure 3.3.26. When using the period 1955-1998, the reference points $\mathbf{F}_{\text {med }}$ and $\mathbf{F}_{\text {high }}$ are about 0.54 and 0.87 , respectively.

The SG on Precautionary Reference Points for Advice on Fishery Management (SGPRP - February 2003) suggested a candidate for $\mathbf{B}_{\text {lim }}$ "somewhere in the range of 400 kt ". Considering that ACFM is unlikely to define and use new $\mathbf{B}_{\text {lim }}$ points, the Working Group will consider the issue further during its 2004 meeting.

The inclusion of the stock recruitment relationship has a major effect on long-term predictions. From Figure 3.3.26 it is seen that below-median recruitment occurs more frequently when the SSB is below-median than when the SSB is above the median. The increased probability of poor recruitment at low SSB is of major concern.

### 3.3.7 Medium-term simulation

The AD-CAM model was used for medium-term simulations using the following premises:

- The amended Harvest Control law was followed.
- Assessment error was assumed to be lognormal with CV of $15 \%$ and autocorrelation 0.2.
- Deviations in weights-at-age were assumed to be lognormal with CV 0.1 and autocorrelation 0.35 . The same deviations were applied to all age groups in the same year. The values are based on examination of weight-at-age in the catches 1980-2002. Errors in weights-at-age and assessment errors were not correlated but it is likely that sudden reduction in weight-at-age will not be predicted and lead to too high catches.

The results of the simulations are shown in figure 3.3.27. The results indicate low probability that the catchable biomass will at the low level observed in the last decade.

### 3.3.8 Management considerations

Catch quotas for the Icelandic cod stock have since 1994 been based on the $25 \%$ catch rule. This catch rule was based on extensive simulations and has been considered precautionary. Until year 2000 the Icelandic government followed the catch rule with minimal deviations although it has turned out that the TAC has exceeded the $25 \%$ rule due to overestimation of the stock. In 2000 the Icelandic government, after some limited studies by the MRI, changed the adopted $25 \%$ catchrule by limiting the allowed changes in TAC between years to 30 thousand tonnes. The catch control rule has been in a reviewing process since 2001 by a group scientists appointed by the Ministry of Fisheries. This group is supposed to deliver a final report in this year.

Since the implementation of the catch rule in 1995 realised reference fishing mortalities have been in the range of 0.550.78 , in last three years about 0.75 . The expected long-term fishing mortality by the application of catch rule was 0.4 .

At present fishing mortality is high (F5-10 in the year 2003 about 0.6 ) and age 6 and younger fish account for more than $86 \%$ of the fishable biomass(4+). This will be reflected in the age composition of the catches in 2004, age group 7 and younger will represent about $88 \%$ of the landings. The age composition of the spawning stock is highly skewed. Spawners at age 6 and younger will constitute to about $70 \%$ of the spawning stock biomass in 2004 and fishes older than ten years old less than $2 \%$. Given the relatively high proportions of younger fish in both the fishable as well as in the spawning stock biomass a lower fishing mortalities than resulting from the catch control rule should be considered.

The working group noticed that before The Ministry of Fisheries allocates the national TAC between vessels, catches of about 2000 tonnes are assumed to be taken by about 300 small jiggers operating in an effort control system. In recent years this amount has been exceeded considerable and in the fishing year 2001/2002 the catches of this fleet was about 12400 t . This is taken into account by the working group estimate of the total annual catch in the assessment year.

### 3.3.9

 Comments on the assessmentThe current assessment and last year assessment are more consistent with previous years assessments compared to the assessments in 1998-2000 were substantial overestimation was observed. As in two previous years assessment indices from commercial fleets were not used for the calibration of the assessment models used. This decision was based on retrospective patterns, the results from the working group on Icelandic cod in autumn 2000 and a study by Guðmundsson and Jónsson (WD-31, 2002) revealing marked trend in catchability in cpue series from commercial fleets. Indices from commercial fleets are still used even if they are not used directly in tuning and they are taken as an important source of information on the state of the stock. The commercial cpue series give the same main message as the survey and a situation where they would show opposite trends would demand thorough investigation of the survey and the cpue indices.

The fishable biomass $4+$ in 2002 was estimated at 680 thous. tonnes in last years assessment compared to 704000 t in the current assessment. This difference of 24000 t, or less than $4 \%$, is well within the error limits of last years point estimate. The SSB is now estimated to have been 357000 t at spawning time in the year 2002. The last years estimate was markedly lower or only 285 000t. Higher observed maturity-at-age in 2002 than assumed for age groups 3-6 does account for the bulk of this increase, but some increase in mean weight-at-age was also observed for age groups less than 10 in the SSB. The relatively high maturity-at-age of younger age groups sampled from the catches at spawning time result in unrealistic high estimate of The SSB especially when strong year classes are entering the fishery. At present a new approach for SSB calculation using maturity and mean weight-at-age from survey is being considered.

The year classes 1998-2001 were estimated $165,175,210$ and 80 millions respectively in last years assessment compared to $165,165,205$ and 70 in the current assessment.

The main causes of the 13-24\% overestimation of this stock in the years 1998-2000 is now considered to be the use of combination of commercial cpue and survey indices for calibration of stock assessment models and high availability of cod in the years 1997 and 1998. The causes for the anticipated increase changes in availability in these years are still not quite understood. Many factors have been mentioned such as: hydrographical changes, capelin availability, increased availability with reduced effort (disturbance), increased natural mortality, emigration, increased discards etc.. Some of those theories have been analysed but no analytical results are available. As those effects still remain unexplained the point estimate in this year's assessment is not corrected for possible changes in parameters of this kind.

Table 3.3.1 Nominal catch (tonnes) of Cod in Division Va, by countries, 1988-2002 as officially reported to ICES.

| Country | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Belgium | 365 | 309 | 260 | 548 | 222 | 145 | 136 | - |
| Faroe Islands | 1,966 | 2,012 | 1,782 | 1,323 | 883 | 664 | 739 |  |
| Germany | - | - | - | - | - | - | - | - |
| Greenland | - | - | - | - | - | - | - | - |
| Iceland | 375,741 | 353,985 | 333,348 | 306,697 | 266,662 | 251,170 | 177,919 | 168,685 |
| Norway | 4 | 3 | - | - | - | - | - | - |
| UK | - | - | - | - | - | - | - | - |
| Total | 378,076 | 356,309 | 335,390 | 308,568 | 267,767 | 251,979 | 178,809 | 169,424 |
| WG estimate | - | - | - | - | - | - | - | - |


| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | $2002^{1}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Belgium | - | - | - | - | - | - | - |
| Faroe Islands | 599 | 408 | 1,078 | 1,247 | 1,176 | 1,129 | 1,188 |
| Germany | - | - | 9 | 21 | 15 | 11 | 15 |
| Greenland | - | - | - | 25 | - | - | - |
| Iceland | 181,052 | 202,745 | 241,545 | 258,658 | 234,362 | 233,875 | 206,745 |
| Norway | 7 | - | - | 85 | 60 | 129 | 76 |
| UK | - | - | - | 16 | 10 | 20 | 32 |
| Total | 181,658 | 203,153 | 242,632 | 260,052 | 235,623 | 235,164 | 208,056 |
| WG estimate | - | - | - | - | - | - | 208,830 |
| 1) Provisional. |  |  |  |  |  |  |  |

Table 3.3.2 Cod at Iceland. Division Va. Landings (tonnes), effort, cpue and percentage changes in effort and cpue in the period 1991-2002 (with 1991 as 100\%). Data are based on logbooks which have been mandatory in the fisheries since 1991.

Bottom trawl

| Year | Catch | Effort | Effort \% changes | Cpue | Cpue \% changes |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 175142 | 234946 | 100 | 745 | 100 |
| 1992 | 131504 | 228196 | 97 | 576 | 77 |
| 1993 | 114587 | 182882 | 78 | 627 | 84 |
| 1994 | 66186 | 83975 | 36 | 788 | 106 |
| 1995 | 60580 | 71202 | 30 | 851 | 114 |
| 1996 | 66867 | 66867 | 28 | 1000 | 134 |
| 1997 | 81202 | 74841 | 32 | 1085 | 146 |
| 1998 | 109947 | 86098 | 37 | 1277 | 171 |
| 1999 | 124381 | 120408 | 51 | 1033 | 139 |
| 2000 | 103289 | 126270 | 54 | 818 | 110 |
| 2001 | 98067 | 109877 | 47 | 892 | 120 |
| 2002 | 88059 | 84340 | 36 | 1044 | 140 |

Gillnet

| Gilinet |  |  |  |  |  |
| :--- | ---: | ---: | :---: | ---: | :---: |
| Year | Catch | Effort | Effort <br> \% changes | Cpue <br> Cpue | changes |
| $\mathbf{1 9 9 1}$ | 58948 | 1060 | 100 | 56 | 100 |
| $\mathbf{1 9 9 2}$ | 59712 | 984 | 93 | 61 | 109 |
| $\mathbf{1 9 9 3}$ | 56701 | 1008 | 95 | 56 | 101 |
| $\mathbf{1 9 9 4}$ | 39192 | 718 | 68 | 55 | 98 |
| $\mathbf{1 9 9 5}$ | 32309 | 437 | 41 | 74 | 133 |
| $\mathbf{1 9 9 6}$ | 41764 | 492 | 46 | 85 | 153 |
| $\mathbf{1 9 9 7}$ | 46742 | 483 | 46 | 97 | 174 |
| $\mathbf{1 9 9 8}$ | 51554 | 721 | 68 | 72 | 129 |
| $\mathbf{1 9 9 9}$ | 47648 | 781 | 74 | 61 | 110 |
| $\mathbf{2 0 0 0}$ | 47989 | 842 | 79 | 57 | 102 |
| $\mathbf{2 0 0 1}$ | 53943 | 1124 | 106 | 48 | 86 |
| $\mathbf{2 0 0 2}$ | 44560 | 990 | 93 | 45 | 81 |

Long line

| Year | Catch | Effort | Effort \% changes | Cpue | Cpue \% changes |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 44711 | 2006 | 100 | 22 | 100 |
| 1992 | 42301 | 2016 | 100 | 21 | 94 |
| 1993 | 47263 | 2224 | 111 | 21 | 95 |
| 1994 | 36426 | 1652 | 82 | 22 | 99 |
| 1995 | 44588 | 1724 | 86 | 26 | 116 |
| 1996 | 39770 | 1478 | 74 | 27 | 121 |
| 1997 | 31276 | 824 | 41 | 38 | 170 |
| 1998 | 37243 | 972 | 48 | 38 | 172 |
| 1999 | 53380 | 1570 | 78 | 34 | 153 |
| 2000 | 50085 | 1727 | 86 | 29 | 130 |
| 2001 | 47092 | 1811 | 90 | 26 | 117 |
| 2002 | 42155 | 1405 | 70 | 30 | 135 |

Table 3.3.3 Cod at Iceland. Catch in numbers by year and age (millions).

| Year/age | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | :--- | :--- | :--- | :--- | :--- |
| $\mathbf{1 9 8 3}$ | 3.554 | 10.910 | 24.305 | 18.944 | 17.382 | 8.381 | $\mathbf{2 . 0 5 4}$ | 2.733 | 0.514 | 0.215 | 0.064 |
| $\mathbf{1 9 8 4}$ | 6.750 | 31.553 | 19.420 | 15.326 | 8.082 | 7.336 | 2.680 | 0.512 | 0.538 | 0.195 | 0.090 |
| $\mathbf{1 9 8 5}$ | 6.457 | 24.552 | 35.392 | 18.267 | 8.711 | 4.201 | 2.264 | 1.063 | 0.217 | 0.233 | 0.102 |
| $\mathbf{1 9 8 6}$ | 20.642 | 20.330 | 26.644 | 30.839 | 11.413 | 4.441 | 1.771 | 0.805 | 0.392 | 0.103 | 0.076 |
| $\mathbf{1 9 8 7}$ | 11.002 | 62.130 | 27.192 | 15.127 | 15.695 | 4.159 | 1.463 | 0.592 | 0.253 | 0.142 | 0.046 |
| $\mathbf{1 9 8 8}$ | 6.713 | 39.323 | 55.895 | 18.663 | 6.399 | 5.877 | 1.345 | 0.455 | 0.305 | 0.157 | 0.114 |
| $\mathbf{1 9 8 9}$ | 2.605 | 27.983 | 50.059 | 31.455 | 6.010 | 1.915 | 0.881 | 0.225 | 0.107 | 0.086 | 0.038 |
| $\mathbf{1 9 9 0}$ | 5.785 | 12.313 | 27.179 | 44.534 | 17.037 | 2.573 | 0.609 | 0.322 | 0.118 | 0.050 | 0.015 |
| $\mathbf{1 9 9 1}$ | 8.554 | 25.131 | 15.491 | 21.514 | 25.038 | 6.364 | 0.903 | 0.243 | 0.125 | 0.063 | 0.011 |
| $\mathbf{1 9 9 2}$ | 12.217 | 21.708 | 26.524 | 11.413 | 10.073 | 8.304 | 2.006 | 0.257 | 0.046 | 0.032 | 0.012 |
| $\mathbf{1 9 9 3}$ | 20.500 | 33.078 | 15.195 | 13.281 | 3.583 | 2.785 | 2.707 | 1.181 | 0.180 | 0.034 | 0.011 |
| $\mathbf{1 9 9 4}$ | 6.160 | 24.142 | 19.666 | 6.968 | 4.393 | 1.257 | 0.599 | 0.508 | 0.283 | 0.049 | 0.018 |
| $\mathbf{1 9 9 5}$ | 10.770 | 9.103 | 16.829 | 13.066 | 4.115 | 1.596 | 0.313 | 0.184 | 0.156 | 0.141 | 0.029 |
| $\mathbf{1 9 9 6}$ | 5.356 | 14.886 | 7.372 | 12.307 | 9.430 | 2.157 | 0.837 | 0.208 | 0.076 | 0.065 | 0.055 |
| $\mathbf{1 9 9 7}$ | 1.722 | 16.442 | 17.298 | 6.711 | 7.379 | 5.958 | 1.147 | 0.493 | 0.126 | 0.028 | 0.037 |
| $\mathbf{1 9 9 8}$ | 3.548 | 7.707 | 25.394 | 20.167 | 5.893 | 3.856 | 2.951 | 0.500 | 0.196 | 0.055 | 0.033 |
| $\mathbf{1 9 9 9}$ | 2.525 | 19.554 | 15.226 | 24.622 | 12.966 | 2.795 | 1.489 | 0.748 | 0.140 | 0.046 | 0.010 |
| $\mathbf{2 0 0 0}$ | 10.493 | 6.581 | 29.080 | 11.227 | 11.390 | 5.714 | 1.104 | 0.567 | 0.314 | 0.074 | 0.022 |
| $\mathbf{2 0 0 1}$ | 11.338 | 25.040 | 9.311 | 19.471 | 5.620 | 3.929 | 2.017 | 0.452 | 0.202 | 0.118 | 0.013 |
| $\mathbf{2 0 0 2}$ | 5.934 | 18.482 | 24.297 | 6.874 | 8.943 | 2.227 | 1.353 | 0.689 | 0.123 | 0.040 | 0.041 |

Table 3.3.4 Cod at Iceland. Division Va. Proportion of fishing and natural mortality before spawning

| $\mathbf{F}$ | $\mathbf{M}$ |
| ---: | ---: |
| 0.085 | 0.250 |
| 0.180 | 0.250 |
| 0.248 | 0.250 |
| 0.296 | 0.250 |
| 0.382 | 0.250 |
| 0.437 | 0.250 |
| 0.477 | 0.250 |
| 0.477 | 0.250 |
| 0.477 | 0.250 |
| 0.477 | 0.250 |
| 0.477 | 0.250 |
| 0.477 | 0.250 |

Table 3.3.5 Cod at Iceland. Division Va. Mean weight-at-age in the landings $(\mathrm{g})$.

| Year/age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1983 | 1095 | 1599 | 2275 | 3021 | 4096 | 5481 | 7049 | 8128 | 11009 | 13972 | 15882 | 18498 |
| 1984 | 1288 | 1725 | 2596 | 3581 | 4371 | 5798 | 7456 | 9851 | 11052 | 14338 | 15273 | 16660 |
| 1985 | 1407 | 1971 | 2576 | 3650 | 4976 | 6372 | 8207 | 10320 | 12197 | 14683 | 16175 | 19050 |
| 1986 | 1459 | 1961 | 2844 | 3593 | 4635 | 6155 | 7503 | 9084 | 10356 | 15283 | 14540 | 15017 |
| 1987 | 1316 | 1956 | 2686 | 3894 | 4716 | 6257 | 7368 | 9243 | 10697 | 10622 | 15894 | 12592 |
| 1988 | 1438 | 1805 | 2576 | 3519 | 4930 | 6001 | 7144 | 8822 | 9977 | 11732 | 14156 | 13042 |
| 1989 | 1186 | 1813 | 2590 | 3915 | 5210 | 6892 | 8035 | 9831 | 11986 | 10003 | 12611 | 16045 |
| 1990 | 1290 | 1704 | 2383 | 3034 | 4624 | 6521 | 8888 | 10592 | 10993 | 14570 | 15732 | 17290 |
| 1991 | 1309 | 1899 | 2475 | 3159 | 3792 | 5680 | 7242 | 9804 | 9754 | 14344 | 14172 | 20200 |
| 1992 | 1289 | 1768 | 2469 | 3292 | 4394 | 5582 | 6830 | 8127 | 12679 | 13410 | 15715 | 11267 |
| 1993 | 1392 | 1887 | 2772 | 3762 | 4930 | 6054 | 7450 | 8641 | 10901 | 12517 | 14742 | 16874 |
| 1994 | 1443 | 2063 | 2562 | 3659 | 5117 | 6262 | 7719 | 8896 | 10847 | 12874 | 14742 | 17470 |
| 1995 | 1348 | 1959 | 2920 | 3625 | 5176 | 6416 | 7916 | 10273 | 11022 | 11407 | 13098 | 15182 |
| 1996 | 1457 | 1930 | 3132 | 4141 | 4922 | 6009 | 7406 | 9772 | 10539 | 13503 | 13689 | 16194 |
| 1997 | 1484 | 1877 | 2878 | 4028 | 5402 | 6386 | 7344 | 8537 | 10797 | 11533 | 10428 | 12788 |
| 1998 | 1230 | 1788 | 2477 | 3588 | 5013 | 7293 | 7843 | 9283 | 10976 | 15352 | 17718 | 16068 |
| 1999 | 1241 | 1716 | 2426 | 3443 | 4720 | 6352 | 8730 | 9946 | 11088 | 12535 | 14995 | 15151 |
| 2000 | 1308 | 1782 | 2330 | 3252 | 4690 | 5894 | 7809 | 9203 | 10240 | 11172 | 13172 | 17442 |
| 2001 | 1499 | 2050 | 2649 | 3413 | 4766 | 6508 | 7520 | 9055 | 8796 | 9526 | 11210 | 13874 |
| 2002 | 1294 | 1926 | 2656 | 3680 | 4720 | 6369 | 7808 | 9002 | 10422 | 13402 | 9008 | 16893 |

Table 3.3.6 Cod at Iceland. Division Va. Mean weight-at-age in the spawning stock(g)

| Year/age | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ |
| ---: | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $\mathbf{1 9 8 3}$ | 891 | 1472 | 2139 | 2918 | 4130 | 5553 | $\mathbf{7 0 0 7}$ | $\mathbf{7 7 7 0}$ | 10817 | 13176 | 14175 |
| $\mathbf{1 9 8 4}$ | 1002 | 1479 | 2257 | 3476 | 4480 | 5887 | 7660 | 9920 | 11035 | 14531 | 15378 |
| $\mathbf{1 9 8 5}$ | 1131 | 1597 | 2285 | 3524 | 5010 | 6195 | 7800 | 9225 | 11336 | 13277 | 15325 |
| $\mathbf{1 9 8 6}$ | 1182 | 1762 | 2681 | 3562 | 4824 | 6457 | 7843 | 9419 | 10674 | 13660 | 13812 |
| $\mathbf{1 9 8 7}$ | 1289 | 1811 | 2735 | 4202 | 5110 | 6497 | 7802 | 10220 | 11197 | 10620 | 15893 |
| $\mathbf{1 9 8 8}$ | 1218 | 1604 | 2499 | 3566 | 5161 | 6238 | 7302 | 8647 | 10184 | 11504 | 14159 |
| $\mathbf{1 9 8 9}$ | 1012 | 1542 | 2423 | 3743 | 5298 | 6910 | 7725 | 9397 | 11953 | 9529 | 12195 |
| $\mathbf{1 9 9 0}$ | 813 | 1330 | 2132 | 3187 | 4691 | 6627 | 8915 | 10362 | 12093 | 15453 | 15337 |
| $\mathbf{1 9 9 1}$ | 1122 | 1776 | 2233 | 3044 | 3891 | 5897 | 7657 | 10573 | 11230 | 14340 | 14172 |
| $\mathbf{1 9 9 2}$ | 876 | 1389 | 2174 | 3185 | 4481 | 5587 | 6775 | 8225 | 11702 | 13474 | 15436 |
| $\mathbf{1 9 9 3}$ | 1037 | 1570 | 2518 | 3611 | 4872 | 6150 | 7538 | 8840 | 11088 | 12002 | 14402 |
| $\mathbf{1 9 9 4}$ | 1193 | 1748 | 2382 | 3684 | 5175 | 6210 | 7676 | 8814 | 10842 | 12595 | 14402 |
| $\mathbf{1 9 9 5}$ | 1066 | 1826 | 2735 | 3497 | 4741 | 6126 | 7582 | 9887 | 10829 | 11307 | 13098 |
| $\mathbf{1 9 9 6}$ | 1264 | 1627 | 2600 | 3829 | 4605 | 5792 | 7550 | 9433 | 11293 | 12984 | 13821 |
| $\mathbf{1 9 9 7}$ | 1221 | 1613 | 2595 | 3807 | 5434 | 6440 | 7629 | 8606 | 10486 | 11774 | 10943 |
| $\mathbf{1 9 9 8}$ | 1260 | 2018 | 2335 | 3529 | 5321 | 7731 | 8173 | 9397 | 10995 | 15274 |  |
| $\mathbf{1 9 9 9}$ | 1068 | 1459 | 2231 | 3181 | 4743 | 6577 | 8561 | 10081 | 11200 | 12567 | 14995 |
| $\mathbf{2 0 0 0}$ | 1025 | 1498 | 2159 | 3236 | 4655 | 5957 | 7881 | 9458 | 10231 | 11736 | 13172 |
| $\mathbf{2 0 0 1}$ | 1121 | 1621 | 2417 | 3234 | 4854 | 6546 | 7935 | 9196 | 9086 | 9899 | 10351 |
| $\mathbf{2 0 0 2}$ | 1004 | 1701 | 2464 | 3673 | 4982 | 6780 | 8328 | 9328 | 10789 | 13983 | 14457 |
| 16894 |  |  |  |  |  |  |  |  |  |  |  |

Table 3.3.7 Cod at Iceland. Division Va. Maturity-at-age in the SSB.

| Year/age | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 8 2}$ | 0.02 | 0.05 | 0.13 | 0.23 | 0.54 | 0.85 | 0.96 | 0.97 | 1.00 | 1.00 | 1.00 |
| $\mathbf{1 9 8 3}$ | 0.00 | 0.09 | 0.17 | 0.34 | 0.51 | 0.72 | 0.86 | 0.98 | 0.98 | 1.00 | 1.00 |
| $\mathbf{1 9 8 4}$ | 0.00 | 0.04 | 0.19 | 0.42 | 0.66 | 0.78 | 0.86 | 0.95 | 0.97 | 0.95 | 1.00 |
| $\mathbf{1 9 8 5}$ | 0.03 | 0.06 | 0.20 | 0.55 | 0.77 | 0.90 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |
| $\mathbf{1 9 8 6}$ | 0.01 | 0.05 | 0.24 | 0.54 | 0.76 | 0.89 | 0.98 | 0.96 | 0.99 | 1.00 | 1.00 |
| $\mathbf{1 9 8 7}$ | 0.02 | 0.05 | 0.24 | 0.59 | 0.81 | 0.94 | 0.95 | 1.00 | 0.98 | 1.00 | 1.00 |
| $\mathbf{1 9 8 8}$ | 0.04 | 0.02 | 0.21 | 0.48 | 0.69 | 0.83 | 0.93 | 0.95 | 0.97 | 0.82 | 1.00 |
| $\mathbf{1 9 8 9}$ | 0.00 | 0.05 | 0.23 | 0.55 | 0.82 | 0.86 | 0.89 | 0.99 | 1.00 | 0.90 | 0.86 |
| $\mathbf{1 9 9 0}$ | 0.00 | 0.08 | 0.30 | 0.63 | 0.82 | 0.91 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 |
| $\mathbf{1 9 9 1}$ | 0.00 | 0.06 | 0.21 | 0.54 | 0.78 | 0.89 | 0.95 | 0.84 | 1.00 | 1.00 | 1.00 |
| $\mathbf{1 9 9 2}$ | 0.07 | 0.23 | 0.56 | 0.71 | 0.91 | 0.96 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 |
| $\mathbf{1 9 9 3}$ | 0.08 | 0.25 | 0.47 | 0.71 | 0.94 | 0.98 | 0.97 | 0.97 | 1.00 | 1.00 | 1.00 |
| $\mathbf{1 9 9 4}$ | 0.10 | 0.28 | 0.57 | 0.80 | 0.90 | 0.92 | 1.00 | 0.85 | 0.99 | 1.00 | 1.00 |
| $\mathbf{1 9 9 5}$ | 0.04 | 0.39 | 0.73 | 0.85 | 0.85 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| $\mathbf{1 9 9 6}$ | 0.08 | 0.10 | 0.51 | 0.74 | 0.86 | 0.91 | 0.84 | 1.00 | 1.00 | 0.99 | 0.97 |
| $\mathbf{1 9 9 7}$ | 0.07 | 0.31 | 0.50 | 0.74 | 0.88 | 0.92 | 0.97 | 0.93 | 1.00 | 0.91 | 1.00 |
| $\mathbf{1 9 9 8}$ | 0.03 | 0.26 | 0.48 | 0.65 | 0.83 | 0.94 | 0.99 | 0.93 | 1.00 | 1.00 | 1.00 |
| $\mathbf{1 9 9 9}$ | 0.05 | 0.30 | 0.55 | 0.72 | 0.83 | 0.93 | 0.97 | 0.99 | 1.00 | 1.00 | 0.84 |
| $\mathbf{2 0 0 0}$ | 0.04 | 0.18 | 0.44 | 0.64 | 0.80 | 0.92 | 0.98 | 0.98 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 0 1}$ | 0.13 | 0.41 | 0.61 | 0.79 | 0.92 | 0.90 | 0.97 | 0.99 | 1.00 | 1.00 | 1.00 |
| $\mathbf{2 0 0 2}$ | 0.12 | 0.41 | 0.59 | 0.84 | 0.85 | 0.99 | 0.99 | 0.98 | 1.00 | 1.00 | 1.00 |

Table 3.3.8 CPUE from bottom trawl survey 1985-2003 as used in the XSA tuning. Sum of North and South (stratified mean) areas indices.

| Year/age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 16.54 | 111.07 | 34.85 | 48.09 | 64.30 | 22.57 | 14.86 | 4.85 | 3.21 | 1.52 | 0.30 | 0.30 | 0.10 |
| 1986 | 15.08 | 60.56 | 95.56 | 22.43 | 21.23 | 26.36 | 6.64 | 2.48 | 0.83 | 0.74 | 0.27 | 0.07 | 0.06 |
| 1987 | 3.65 | 28.86 | 103.10 | 82.03 | 21.08 | 12.22 | 12.02 | 2.57 | 0.90 | 0.40 | 0.45 | 0.23 | 0.15 |
| 1988 | 3.44 | 7.36 | 71.69 | 101.61 | 66.75 | 7.81 | 5.88 | 6.14 | 0.58 | 0.25 | 0.11 | 0.12 | 0.05 |
| 1989 | 4.04 | 16.45 | 21.97 | 77.70 | 67.59 | 34.20 | 4.20 | 1.45 | 1.14 | 0.24 | 0.17 | 0.06 | 0.01 |
| 1990 | 5.56 | 11.79 | 26.15 | 14.07 | 26.97 | 32.38 | 14.22 | 1.51 | 0.53 | 0.42 | 0.13 | 0.00 | 0.04 |
| 1991 | 3.95 | 16.27 | 17.93 | 30.17 | 15.24 | 18.09 | 20.93 | 4.23 | 0.80 | 0.32 | 0.24 | 0.00 | 0.11 |
| 1992 | 0.72 | 17.13 | 33.26 | 18.87 | 16.27 | 6.54 | 5.70 | 5.11 | 1.29 | 0.22 | 0.04 | 0.04 | 0.04 |
| 1993 | 3.57 | 4.82 | 30.76 | 36.41 | 13.24 | 9.93 | 2.13 | 1.75 | 1.17 | 0.34 | 0.11 | 0.03 | 0.03 |
| 1994 | 14.38 | 15.01 | 8.97 | 26.66 | 21.90 | 5.77 | 3.62 | 0.70 | 0.48 | 0.43 | 0.14 | 0.02 | 0.03 |
| 1995 | 1.18 | 29.03 | 24.78 | 8.99 | 23.88 | 17.69 | 3.78 | 1.76 | 0.35 | 0.17 | 0.21 | 0.12 | 0.02 |
| 1996 | 3.72 | 5.48 | 42.60 | 29.44 | 12.84 | 14.62 | 13.99 | 3.81 | 1.05 | 0.19 | 0.06 | 0.22 | 0.09 |
| 1997 | 1.21 | 22.39 | 13.57 | 56.18 | 29.05 | 9.48 | 8.71 | 6.59 | 0.56 | 0.36 | 0.15 | 0.04 | 0.12 |
| 1998 | 8.06 | 5.56 | 29.98 | 16.06 | 61.77 | 28.33 | 6.51 | 5.20 | 3.05 | 0.66 | 0.13 | 0.00 | 0.02 |
| 1999 | 7.39 | 32.98 | 7.01 | 42.27 | 13.02 | 23.66 | 11.12 | 2.35 | 1.32 | 0.66 | 0.15 | 0.06 | 0.00 |
| 2000 | 18.79 | 27.90 | 54.74 | 6.94 | 30.00 | 8.28 | 8.18 | 4.14 | 0.51 | 0.30 | 0.07 | 0.03 | 0.04 |
| 2001 | 12.16 | 21.72 | 36.78 | 37.60 | 4.91 | 15.24 | 3.33 | 1.97 | 0.79 | 0.23 | 0.10 | 0.09 | 0.04 |
| 2002 | 0.92 | 38.07 | 41.12 | 40.16 | 36.16 | 7.10 | 8.33 | 1.49 | 0.72 | 0.30 | 0.00 | 0.01 | 0.00 |
| 2003 | 11.17 | 4.44 | 46.36 | 38.55 | 31.51 | 19.09 | 4.11 | 4.71 | 1.08 | 0.23 | 0.09 | 0.02 | 0.06 |

Table 3.3.9
CPUE from bottom trawl survey 1985-2003 as used in the TSA runs. Weighted geometric mean of North and South (stratified mean) areas indices.

| Year/age | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1 9 8 5}$ | 18705 | 19615 | 9483 | 4147 | 1973 | 1369 |
| $\mathbf{1 9 8 6}$ | 9273 | 8933 | 10327 | 2941 | 813 | 381 |
| $\mathbf{1 9 8 7}$ | 31777 | 7042 | 5180 | 4004 | 1163 | 349 |
| $\mathbf{1 9 8 8}$ | 39727 | 25703 | 3673 | 1976 | 1681 | 284 |
| $\mathbf{1 9 8 9}$ | 30658 | 27853 | 16330 | 1989 | 636 | 391 |
| $\mathbf{1 9 9 0}$ | 4033 | 12061 | 16118 | 6125 | 727 | 269 |
| $\mathbf{1 9 9 1}$ | 11804 | 6284 | 8785 | 10237 | 2059 | 384 |
| $\mathbf{1 9 9 2}$ | 8485 | 6575 | 2829 | 2834 | 2772 | 677 |
| $\mathbf{1 9 9 3}$ | 17146 | 6249 | 4572 | 907 | 870 | 552 |
| $\mathbf{1 9 9 4}$ | 13392 | 9484 | 2732 | 1296 | 292 | 235 |
| $\mathbf{1 9 9 5}$ | 4516 | 10885 | 8169 | 1696 | 728 | 166 |
| $\mathbf{1 9 9 6}$ | 14321 | 5733 | 7050 | 6093 | 1525 | 500 |
| $\mathbf{1 9 9 7}$ | 21384 | 11060 | 4014 | 3560 | 2422 | 261 |
| $\mathbf{1 9 9 8}$ | 6127 | 18163 | 10267 | 2393 | 1666 | 1211 |
| $\mathbf{1 9 9 9}$ | 15378 | 5269 | 11090 | 4955 | 1182 | 488 |
| $\mathbf{2 0 0 0}$ | 3106 | 12299 | 3798 | 3392 | 2064 | 255 |
| $\mathbf{2 0 0 1}$ | 16395 | 2157 | 6363 | 1297 | 735 | 393 |
| $\mathbf{2 0 0 2}$ | 18763 | 16991 | 3594 | 3082 | 712 | 311 |
| $\mathbf{2 0 0 3}$ | 14976 | 13062 | 8562 | 2034 | 1133 | 347 |

Table 3.3.10 a XSA Tuning diagnostic. Survey indices agegroups 3-9.

```
Lowestoft VPA Version 3.1
    12/04/2003 18:00
Extended Survivors Analysis
"ICELANDIC COD (Div. Va); data from 1972-2002"
CPUE data from file codvarnt.dat
Catch data for 32 years. 1971 to 2002. Ages 3 to 14.
    Fleet, First, Last, First, Last, Alpha, Beta
```


Time-series weights :
Tapered time weighting not applied
Catchability analysis :
Catchability dependent on stock size for ages < 5
Regression type $=C$
Minimum of 5 points used for regression
Survivor estimates shrunk to the population mean for ages < 5
Catchability independent of age for ages $>=11$
Terminal population estimation :
Survivor estimates shrunk towards the mean $F$
of the final 3 years or the 4 oldest ages.
S.E. of the mean to which the estimates are shrunk $=$. 500
Minimum standard error for population
estimates derived from each fleet $=$. 300
Prior weighting not applied
Tuning converged after 14 iterations
1
Regression weights
, $1.000,1.000,1.000,1.000,1.000,1.000,1.000,1.000,1.000,1.000$
Fishing mortalities
Age, 1993, 1994, 1995, 1996, 1997, 1998, 1999, 2000, 2001, 2002

| 3, | .167, | .095, | .076, | .036, | .022, | .024, | .049, | .068, | .080, | .043 |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 4, | .323, | .304, | .199, | .143, | .148, | .132, | .182, | .175, | .228, | .182 |
| 5, | .501, | .324, | .360, | .246, | .247, | .357, | .416, | .451, | .401, | .361 |
| 6, | .782, | .454, | .372, | .489, | .372, | .508, | .709, | .624, | .626, | .589 |
| 7, | .808, | .652, | .534, | .507, | .621, | .659, | .733, | .876, | .754, | .670 |
| 8, | 1.130, | .761, | .524, | .602, | .713, | .796, | .777, | .872, | .892, | .788 |
| 9, | 1.237, | .800, | .426, | .581, | .768, | .991, | .853, | .837, | .915, | .931 |
| 10, | .951, | .823, | .616, | .564, | .837, | .956, | .744, | .984, | 1.064, | .981 |
| 11, | .904, | .625, | .652, | .561, | .822, | 1.009, | .794, | .835, | 1.306, | .998 |
| 12, | .577, | .671, | .752, | .631, | .413, | 1.141, | .693, | 1.527, | .915, | 1.051 |
| 13, | .567, | .702, | 1.174, | .763, | .945, | 1.331, | .640, | .876, | 1.483, | 1.007 |
| 14, | .759, | .710, | .805, | .636, | .764, | 1.123, | .724, | 1.070, | 1.206, | 1.021 |

## Table 3.3.10.a (Cont'd)

XSA population numbers (Thousands)

| AGE |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | , | 3, |  | 4, | 5, |  | 6 , | 7, | 8, |  | 9, |
| 10, |  | 11, | 12 |  |  |  |  |  |  |  |  |
| 1993 | , | 1.47E+05, | 1.32E+05, | 4.26E+04, | 2.71E+04, | 7.14E+03, | 4.55E+03, | 4.21E+03, | 2.13E+03, | 3.34E+02, | 8.57E+01, |
| 1994 | , | $7.48 \mathrm{E}+04$, | 1.02E+05, | 7.85E+04, | 2.11E+04, | 1.01E+04, | 2.61E+03, | 1.20E+03, | 1.00E+03, | $6.73 \mathrm{E}+02$, | 1.11E+02, |
| 1995 | , | 1.63E+05, | $5.57 \mathrm{E}+04$, | $6.15 \mathrm{E}+04$, | 4.64E+04, | 1.10E+04, | 4.33E+03, | 9.97E+02, | 4.42E+02, | 3.60E+02, | 2.95E+02, |
| 1996 | , | 1.68E+05, | 1.23E+05, | 3.73E+04, | 3.51E+04, | 2.62E+04, | $5.27 \mathrm{E}+03$, | 2.10E+03, | $5.33 \mathrm{E}+02$, | 1.96E+02, | 1.54E+02, |
| 1997 | , | 8.61E+04, | 1.32E+05, | 8.75E+04, | 2.39E+04, | 1.76E+04, | 1.29E+04, | 2.36E+03, | 9.61E+02, | 2.48E+02, | 9.15E+01, |
| 1998 | , | 1.62E+05, | $6.89 \mathrm{E}+04$, | 9.35E+04, | 5.60E+04, | 1.35E+04, | 7.76E+03, | 5.19E+03, | $8.97 \mathrm{E}+02$, | 3.41E+02, | 8.93E+01, |
| 1999 | , | 5.81E+04, | 1.30E+05, | 4.95E+04, | 5.36E+04, | $2.76 \mathrm{E}+04$, | $5.72 \mathrm{E}+03$, | 2.87E+03, | 1.58E+03, | 2.82E+02, | 1.02E+02, |
| 2000 | , | 1.78E+05, | 4.53E+04, | 8.86E+04, | 2.67E+04, | 2.16E+04, | 1.09E+04, | 2.15E+03, | 1.00E+03, | $6.13 \mathrm{E}+02$, | 1.04E+02, |
| 2001 | , | 1.63E+05, | 1.36E+05, | 3.11E+04, | 4.62E+04, | 1.17E+04, | 7.36E+03, | 3.72E+03, | 7.63E+02, | 3.06E+02, | 2.18E+02, |
| 2002 | , | 1.54E+05, | 1.23E+05, | 8.86E+04, | 1.71E+04, | 2.02E+04, | 4.51E+03, | $2.47 \mathrm{E}+03$, | 1.22E+03, | 2.15E+02, | 6.80E+01, |

Estimated population abundance at 1st Jan 2003
$0.00 \mathrm{E}+00,1.21 \mathrm{E}+05, \quad 8.40 \mathrm{E}+04,5.06 \mathrm{E}+04,7.75 \mathrm{E}+03, \quad 8.48 \mathrm{E}+03,1.68 \mathrm{E}+03,7.96 \mathrm{E}+02,3.74 \mathrm{E}+02,6.50 \mathrm{E}+01$,
Taper weighted geometric mean of the VPA populations:
$1.62 \mathrm{E}+05,1.26 \mathrm{E}+05, \quad 8.25 \mathrm{E}+04,4.51 \mathrm{E}+04,2.21 \mathrm{E}+04,9.37 \mathrm{E}+03,3.67 \mathrm{E}+03,1.43 \mathrm{E}+03,5.15 \mathrm{E}+02,1.92 \mathrm{E}+02$,
Standard error of the weighted Log(VPA populations) :
.4313, .4281, .4340, .4734, .5278, .6514, .7516, .8145, .7490, .7016,

| YEAR , | 13, |
| :---: | :---: |
| 1993, | $2.81 \mathrm{E}+01$, |
| 1994, | $2.70 \mathrm{E}+01$, |
| 1995, | $4.64 \mathrm{E}+01$, |
| 1996, | $1.30 \mathrm{E}+01$, |
| 1997, | $6.69 \mathrm{E}+02$, |
| 1998 | $1.17 \mathrm{E}+01$, |
| 1999, | $4.96 \mathrm{E}+01$, |
| 2000, | $2.34 \mathrm{E}+01$, |
| 2001, | $4.17 \mathrm{E}+01$, |
| 2002, | $1.86 \mathrm{E}+01$, |
|  | $7.14 \mathrm{E}+01$, |
|  |  |
|  |  |

Estimated population abundance at 1st Jan 2003
1.94E+01, 2.13E+01,

Taper weighted geometric mean of the VPA populations:

$$
, \quad 7.15 \mathrm{E}+01,2.33 \mathrm{E}+01
$$

Standard error of the weighted Log(VPA populations) :
1

Log catchability residuals.

| Age | , | 1984, | 1985, | 1986, | 1987, | 1988, | 1989, | 1990, | 1991, | 1992 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | , | . 28 , | -. 25, | -. 24 , | . 06 , | . 39, | . 00 , | . 03, | . 03, | -. 10 |  |
| 4 | , | . 14 , | -.05, | -.14, | -. 10, | . 00, | -. 10, | . 18, | -.17, | . 05 |  |
| 5 | , | . 27 , | . 04 , | . 07 , | -. 54, | . 18, | -.18, | . 10, | -. 19, | -. 01 |  |
| 6 | , | . 62, | -. 29 , | . 04 , | . 22, | -. 23, | . 00 , | . 12, | -.19, | -. 33 |  |
| 7 | , | . 28 , | -. 29, | -. 15, | . 55, | -. 03, | -. 33, | -. 23, | -. 14, | -. 07 |  |
| 8 | , | . 65, | -.45, | -. 06 , | -. 34, | . 57, | . 15, | . 16, | -. 32, | -. 28 |  |
| Age | , | 1993, | 1994, | 1995, | 1996, | 1997, | 1998, | 1999, | 2000, | 2001, | 2002 |
| 3 | , | -.07, | -.14, | -.16, | . 20, | . 05 , | . 04 , | -.08, | -.10, | . 04 , | . 04 |
| 4 | , | -. 20, | .11, | . 21, | -.05, | . 40 , | -.04, | -. 06 , | -.27, | . 06 , | . 03 |
| 5 | , | -. 30, | . 03 , | . 12 , | . 07 , | . 31 , | . 18, | -. 18, | -.12, | . 12 , | . 02 |
| 6 |  | -. 14, | -.18, | . 26 , | . 18, | . 16, | -. 02, | -.08, | -. 37, | . 00 , | . 25 |
| 7 |  | -. 43, | -. 01, | . 57, | . 22 , | . 49, | . 00 , | -.07, | -. 43, | -. 22, | . 30 |
| 8 | , | . 08 , | -. 05 , | . 31 , | -. 44 , | . 47 , | . 22 , | -. 44, | -. 55, | -. 23 , | . 56 |

## Table 3.3.10.a (Cont'd)

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age , | 5, | 6, | 7, | 8 |
| :---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -7.9049, | -7.7982, | -7.8016, | -7.9110, |
| S.E (Log q), | .2092, | .2482, | .3151, | .3897, |

Regression statistics :
Ages with $q$ dependent on year class strength

| Age, Slope, | t-value, | Intercept, RSquare, | No Pts, Reg s.e, Mean Log |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3, | .64, | 3.840, | 9.48, | .87, | 19, | .17, |
| 4, | .69, | 3.475, | 9.09, | .88, | 19, | -17, |

Ages with $q$ independent of year class strength and constant w.r.t. time.
Age, Slope, t-value, Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 5, | .92, | .779, | 8.17, | .84, | 19, | .19, | -7.90, |
| :--- | ---: | :--- | :--- | :--- | :--- | :--- | :--- |
| 6, | .94, | .512, | 7.97, | .81, | 19, | .24, | -7.80, |
| 7, | .93, | .433, | 7.93, | .72, | 19, | .30, | -7.80, |
| 8, | 1.06, | -.288, | 7.86, | .58, | 19, | .42, | -7.91, |



Regression statistics :
Ages with $q$ dependent on year class strength
Age, Slope, t-value, Intercept, RSquare, No Pts, Reg s.e, Mean Log q


1

Terminal year survivor and $F$ summaries :
Age 3 Catchability dependent on age and year class strength
Year class $=1999$

| Fleet, |  | Estimated, | Int, | Ext, | Var, |  | Scaled, | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Survivors, | s.e, | s.e, | Ratio, | , | Weights, | F |
| SMB. Tot | , | 125999., | . 300, | . 000 , | . 00 , | 1, | . 346 , | . 042 |
| SMB. Tot a3 on a3 | , | 133861., | . 300 , | . 000 , | . 00 , | 1, | . 346 , | . 039 |
| $P$ shrinkage mean | , | 125792., | .43, , , r |  |  |  | . 178, | . 042 |
| F shrinkage mean | , | 78983., | . 50, , , , |  |  |  | . 130, | . 066 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | ---: | ---: | ---: | ---: | ---: |
| at end of year, | s.e, | s.e, | Ratio, |  |  |
| $121044 .$, | .18, | .10, | 4, | .590, | .043 |

1
Age 4 Catchability dependent on age and year class strength
Year class $=1998$


Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year, | s.e, | S.e, | R | Ratio, |  |
| $84030 .$, | .15, | .02, | 5, | .136, | .182 |

Age 5 Catchability constant w.r.t. time and dependent on age
Year class $=1997$

| Fleet, |  | Estimated, | Int, | Ext, | Var, | N, | Scaled, | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Survivors, | s.e, | s.e, | Ratio, |  | Weights, | F |
| SMB. Tot | , | 50423., | . 175, | . 046 , | . 26 , | 3, | . 668, | 362 |
| SMB. Tot a3 on a3 | , | 58755., | . 300 , | .000, | . 00 , | 1, | .196, | 318 |
| F shrinkage mean |  | 41457., | . 50, |  |  |  | .136, | . 425 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | Ratio, |  |  |
| $50592 .$, | .15, | .06, | 5, | .395, | .361 |

1
Age 6 Catchability constant w.r.t. time and dependent on age
Year class $=1996$

| Fleet, |  | Estimated, Survivors, | Int, | Ext, s.e, | Var, Ratio, | N, | Scaled, Weights, | $\underset{\mathrm{F}}{\text { Estimated }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. Tot | , | 8144., | .155, | .115, | . 74 , | 4, | . 700 , | . 567 |
| SMB. Tot a3 on a3 | , | 7224., | . 300 , | . 000 , | . 00 , | 1, | . 135, | 621 |
| F shrinkage mean |  | 6665 | 50 |  |  |  | 164 | 65 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | S.e, | Ratio, |  |  |
| $7754 .$, | .14, | .08, | 6, | .586, | .589 |

Age 7 Catchability constant w.r.t. time and dependent on age
Year class $=1995$

| Fleet, |  | Estimated, | Int, | Ext, | Var, | N, | Scaled, | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Survivors, | s.e, | s.e, | Ratio, |  | Weights, | F |
| SMB. Tot | , | 9302., | .156, | .084, | . 54, | 5, | . 701 , | . 626 |
| SMB. Tot a3 on a3 | , | 7238. | . 300, | .000, | . 00 , | 1 , | . 085, | 750 |
| F shrinkage mean |  | 6655. | . 50, |  |  |  | . 215, | .796 |

Weighted prediction :
Survivors, Int, Ext, N, Var, F


## Table 3.3.10.a (Cont'd)

Age 8 Catchability constant w.r.t. time and dependent on age
Year class $=1994$

| Fleet, |  | Estimated, Survivors, | Int, | Ext, s.e, | Var, <br> Ratio, |  | Scaled, Weights, | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. 'Tot |  | 1786., | .169, | .165, | $\begin{aligned} & \text { Clo, } \\ & .98 \text {, } \end{aligned}$ | 6, | .642, | . 755 |
| SMB. Tot a3 on a3 | , | 1618., | . 300 , | . 000 , | . 00 , | 1, | . 055 , | . 809 |
| $F$ shrinkage mean |  | 1489., | . 50, |  |  |  | . 303, | . 856 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | r, | Ratio, |  |
| $1681 .$, | .19, | .12, | 8, | .630, | .788 |

Age 9 Catchability constant w.r.t. time and dependent on age
Year class $=1993$

| Fleet, |  | Estimated, Survivors, | $\begin{aligned} & \text { Int, } \\ & \text { s.e, } \end{aligned}$ | $\begin{aligned} & \text { Ext, } \\ & \text { s.e, } \end{aligned}$ | Var, <br> Ratio, |  | Scaled, Weights, | $\begin{aligned} & \text { Estimated } \\ & \mathrm{F} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. Tot | , | 696. | . 178 , | 115, | .65, | 6, | . 392 , | 1.015 |
| SMB. Tot a3 on a3 | , | 830. | . 300 , | . 000 , | . 00 , | 1, | . 031, | . 907 |
| F shrinkage mean |  | 871. | . 50, |  |  |  | . 577 , | . 878 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| $796 .$, | .30, | .09, | 8, | .292, | .931 |

1
Age 10 Catchability constant w.r.t. time and dependent on age
Year class $=1992$

| Fleet, |  | Estimated, Survivors, | Int, | Ext, | Var, <br> Ratio, | N, | Scaled, Weights, | Estimated <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. Tot |  | Survivors, | s.e, 158, | $.126,$ | Ratio, | 6 | Weights, | 1.090 |
| SMB. Tot a3 on a3 | , | 284 | . 300, | . 000 , | . 00 , | 1, | . 023, | 1.161 |
| F shrinkage mean |  | 399., | . 50, |  |  |  | . 736 , | . 941 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| $374 .$, | .37, | .09, | 8, | .255, | .981 |

Age 11 Catchability constant w.r.t. time and dependent on age
Year class $=1991$

| Fleet, |  | Estimated, Survivors, | Int, | Ext, <br> s.e, | Var, <br> Ratio, | N, | Scaled, Weights, | $\begin{gathered} \text { Estimated } \\ \mathrm{F} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. Tot | , | 60., | . 151, | .109, | . 72 , | 6, | . 126 , | 1.044 |
| SMB. Tot a3 on a3 | , | $56 .$, | . 300 , | . 000 , | . 00 , | 1, | . 013, | 1.099 |
| F shrinkage mean |  | 66., | 50, |  |  |  | . 861 , | . 989 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| $65 .$, | .43, | .05, | 8, | .108, | .998 |

## Table 3.3.10.a (Cont'd)

Age 12 Catchability constant w.r.t. time and age (fixed at the value for age) 11
Year class $=1990$

| Fleet, |  | Estimated, Survivors, | Int, s.e, | Ext, s.e, | Var, Ratio, | N, | Scaled, Weights, | $\begin{gathered} \text { Estimated } \\ \mathrm{F} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. Tot |  | 25., | . 160 , | . 072 , | . 45, | 6, | . 035, | . 904 |
| SMB. Tot a3 on a3 |  | 19., | . 300 , | . 000 , | . 00 , | 1, | . 003 , | 1.066 |
| F shrinkage mean |  | 19., | . 50, |  |  |  | . 963 , | 1.057 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | r, | Ratio, |  |
| $19 .$, | .48, | .08, | 8, | .176, | 1.051 |

Age 13 Catchability constant w.r.t. time and age (fixed at the value for age) 11
Year class $=1989$

| Fleet, |  | Estimated, Survivors, | Int, s.e, | Ext, s.e, | Var, <br> Ratio, | N, | Scaled, Weights, | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. Tot | , | 26., | . 150, | . 096 , | .64, | 6, | .033, | . 888 |
| SMB. Tot a3 on a3 | , | 18., | . 300 , | . 000 , | . 00 , | 1, | . 003 , | 1.111 |
| F shrinkage mean |  | 21., | . 50, |  |  |  | . 965 , | 1.011 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | r, | Ratio, |  |
| $21 .$, | .48, | .07, | 8, | .136, | 1.007 |

1
Age 14 Catchability constant w.r.t. time and age (fixed at the value for age) 11
Year class $=1988$

| Fleet, |  | Estimated, Survivors, | Int, | Ext, s.e, | Var, <br> Ratio, | N, | Scaled, Weights, | Estimated <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. Tot |  | Survivors | s.e, 160, | s.e, .175, | Ratio, $1.09$ | 6, | Weights, | F 000 |
| SMB. Tot a3 on a3 |  | 1 | . 300 , | . 000 , | . 00 , | 1 , | . 000 , | . 000 |
| F shrinkage mean |  | 1. | . 50, |  |  |  | . 995 , | 1.021 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | Far |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| $1 .$, | .50, | .02, | 8, | .038, | 1.021 |

Table 3.3.10b XSA Tuning diagnostic. Survey indices agegroups 1-9.

```
Lowestoft VPA Version 3.1
    14/04/2003 9:26
Extended Survivors Analysis
"ICELANDIC COD (Div. Va); data from 1972-2002"
CPUE data from file codvarnt.dat
Catch data for 32 years. 1971 to 2002. Ages 0 to 14.
```



```
Time-series weights :
    Tapered time weighting not applied
Catchability analysis :
    Catchability dependent on stock size for ages < 5
    Regression type = C
    Minimum of }5\mathrm{ points used for regression
    Survivor estimates shrunk to the population mean for ages < 5
Catchability independent of age for ages >= 11
Terminal population estimation :
    Survivor estimates shrunk towards the mean F
    of the final 3 years or the 4 oldest ages.
    S.E. of the mean to which the estimates are shrunk = . 500
    Minimum standard error for population
    estimates derived from each fleet = . 300
    Prior weighting not applied
Tuning converged after 14 iterations
```

1
Regression weights
, $1.000,1.000,1.000,1.000,1.000,1.000,1.000,1.000,1.000,1.000$
Fishing mortalities
Age, 1993, 1994, 1995, 1996, 1997, 1998, 1999, 2000, 2001, 2002

| 0, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000 |
| 2, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000, | .000 |
| 3, | .167, | .095, | .076, | .036, | .022, | .024, | .048, | .066, | .078, | .038 |
| 4, | .323, | .304, | .199, | .143, | .148, | .132, | .183, | .169, | .224, | .176 |
| 5, | .501, | .324, | .360, | .246, | .247, | .357, | .417, | .454, | .383, | .352 |
| 6, | .782, | .454, | .372, | .489, | .372, | .508, | .709, | .626, | .635, | .547 |
| 7, | .808, | .652, | .534, | .507, | .621, | .659, | .733, | .874, | .760, | .688 |
| 8, | 1.130, | .761, | .524, | .602, | .713, | .796, | .777, | .873, | .888, | .802 |
| 9, | 1.237, | .800, | .426, | .581, | .768, | .992, | .853, | .838, | .919, | .919 |
| 10, | .951, | .823, | .616, | .564, | .837, | .956, | .744, | .984, | 1.066, | .990 |
| 11, | .904, | .625, | .652, | .561, | .822, | 1.009, | .794, | .836, | 1.306, | 1.000 |
| 12, | .577, | .671, | .752, | .631, | .413, | 1.141, | .693, | 1.527, | .915, | 1.052 |
| 13, | .567, | .702, | 1.174, | .763, | .945, | 1.331, | .640, | .877, | 1.484, | 1.009 |
| 14, | .759, | .710, | .805, | .636, | .764, | 1.123, | .724, | 1.070, | 1.207, | 1.024 |

## Table 3.3.10.b (Cont'd)

XSA population numbers (Thousands)

| YEAR | , | 0 , |  | 1, | 2, |  | 3, | 4, | 5, |  | 6, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7, |  | 8, | 9, |  |  |  |  |  |  |  |  |
| 1993 | , | 3.05E+05, | 2.42E+05, | 9.14E+04, | 1.47E+05, | 1.32E+05, | 4.26E+04, | 2.71E+04, | 7.14E+03, | 4.55E+03, | 4.21E+03, |
| 1994 | , | 1.57E+05, | 2.50E+05, | 1.98E+05, | 7.48E+04, | 1.02E+05, | $7.85 \mathrm{E}+04$, | 2.11E+04, | 1.01E+04, | 2.61E+03, | 1.20E+03, |
| 1995 | , | $2.94 \mathrm{E}+05$, | 1.28E+05, | $2.05 \mathrm{E}+05$, | 1.62E+05, | $5.57 \mathrm{E}+04$, | $6.15 \mathrm{E}+04$, | 4.64E+04, | 1.10E+04, | 4.33E+03, | $9.97 \mathrm{E}+02$, |
| 1996 | , | $1.09 \mathrm{E}+05$, | 2.41E+05, | $1.05 \mathrm{E}+05$, | 1.68E+05, | 1.23E+05, | $3.73 \mathrm{E}+04$, | 3.51E+04, | 2.62E+04, | 5.27E+03, | 2.10E+03, |
| 1997 | , | $3.29 \mathrm{E}+05$, | 8.93E+04, | 1.97E+05, | 8.60E+04, | 1.32E+05, | $8.75 \mathrm{E}+04$, | 2.39E+04, | 1.76E+04, | 1.29E+04, | 2.36E+03, |
| 1998 | , | $3.04 \mathrm{E}+05$, | 2.69E+05, | 7.32E+04, | 1.62E+05, | $6.88 \mathrm{E}+04$, | 9.35E+04, | 5.60E+04, | 1.35E+04, | 7.76E+03, | $5.19 \mathrm{E}+03$, |
| 1999 | , | $3.17 \mathrm{E}+05$, | $2.49 \mathrm{E}+05$, | 2.20E+05, | 5.99E+04, | 1.29E+05, | $4.94 \mathrm{E}+04$, | 5.36E+04, | $2.76 \mathrm{E}+04$, | 5.72E+03, | 2.87E+03, |
| 2000 | , | $3.61 \mathrm{E}+05$, | 2.59E+05, | $2.04 \mathrm{E}+05$, | 1.80E+05, | 4.67E+04, | 8.81E+04, | 2.67E+04, | $2.16 \mathrm{E}+04$, | 1.08E+04, | $2.15 \mathrm{E}+03$, |
| 2001 |  | 1.42E+05, | 2.95E+05, | 2.12E+05, | 1.67E+05, | 1.38E+05, | $3.23 \mathrm{E}+04$, | $4.58 \mathrm{E}+04$, | 1.17E+04, | 7.38E+03, | 3.71E+03, |
| 2002 |  | $3.45 \mathrm{E}+05$, | 1.17E+05, | 2.42E+05, | 1.74E+05, | $1.26 \mathrm{E}+05$, | 9.04E+04, | 1.80E+04, | 1.99E+04, | $4.46 \mathrm{E}+03$, | $2.49 \mathrm{E}+03$, |

Estimated population abundance at 1st Jan 2003
$0.00 \mathrm{E}+00,2.82 \mathrm{E}+05,9.54 \mathrm{E}+04,1.98 \mathrm{E}+05,1.37 \mathrm{E}+05,8.68 \mathrm{E}+04,5.20 \mathrm{E}+04,8.55 \mathrm{E}+03,8.17 \mathrm{E}+03,1.64 \mathrm{E}+03$,
Taper weighted geometric mean of the VPA populations:
$2.89 \mathrm{E}+05,2.40 \mathrm{E}+05,2.00 \mathrm{E}+05,1.63 \mathrm{E}+05,1.26 \mathrm{E}+05,8.26 \mathrm{E}+04,4.52 \mathrm{E}+04,2.21 \mathrm{E}+04,9.36 \mathrm{E}+03,3.67 \mathrm{E}+03$, Standard error of the weighted Log(VPA populations) :
.4350, .4489, .4304, .4293, .4259, .4315, .4699, .5281, .6517, .7514,

|  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| YEAR, | 10, | 11, | 12, | 13, |  |
| 1993, | $2.13 \mathrm{E}+03$, | $3.34 \mathrm{E}+02$, | $8.57 \mathrm{E}+01$, | $2.81 \mathrm{E}+01$, | $2.70 \mathrm{E}+01$, |
| 1994, | $1.00 \mathrm{E}+03$, | $6.73 \mathrm{E}+02$, | $1.11 \mathrm{E}+02$, | $3.94 \mathrm{E}+01$, | $1.30 \mathrm{E}+01$, |
| 1995, | $4.42 \mathrm{E}+02$, | $3.60 \mathrm{E}+02$, | $2.95 \mathrm{E}+02$, | $4.64 \mathrm{E}+01$, | $1.60 \mathrm{E}+01$, |
| 1996, | $5.33 \mathrm{E}+02$, | $1.96 \mathrm{E}+02$, | $1.54 \mathrm{E}+02$, | $1.14 \mathrm{E}+02$, | $1.17 \mathrm{E}+01$, |
| 1997, | $9.61 \mathrm{E}+02$, | $2.48 \mathrm{E}+02$, | $9.15 \mathrm{E}+01$, | $6.69 \mathrm{E}+01$, | $4.35 \mathrm{E}+01$, |
| 1998, | $8.97 \mathrm{E}+02$, | $3.41 \mathrm{E}+02$, | $8.93 \mathrm{E}+01$, | $4.96 \mathrm{E}+01$, | $2.13 \mathrm{E}+01$, |
| 1999, | $1.58 \mathrm{E}+03$, | $2.82 \mathrm{E}+02$, | $1.02 \mathrm{E}+02$, | $2.34 \mathrm{E}+01$, | $1.07 \mathrm{E}+01$, |
| 2000, | $1.00 \mathrm{E}+03$, | $6.13 \mathrm{E}+02$, | $1.04 \mathrm{E}+02$, | $4.16 \mathrm{E}+01$, | $1.01 \mathrm{E}+01$, |
| 2001, | $7.62 \mathrm{E}+02$, | $3.06 \mathrm{E}+02$, | $2.18 \mathrm{E}+02$, | $1.86 \mathrm{E}+01$, | $1.42 \mathrm{E}+01$, |
| 2002, | $1.21 \mathrm{E}+03$, | $2.15 \mathrm{E}+02$, | $6.79 \mathrm{E}+01$, | $7.13 \mathrm{E}+01$, | $3.45 \mathrm{E}+00$, |

Estimated population abundance at 1st Jan 2003
$8.12 \mathrm{E}+02,3.69 \mathrm{E}+02,6.47 \mathrm{E}+01,1.94 \mathrm{E}+01,2.13 \mathrm{E}+01$,
Taper weighted geometric mean of the VPA populations:

```
1.43E+03, 5.15E+02, 1.92E+02, 7.15E+01, 2.33E+01,
```

Standard error of the weighted Log(VPA populations) :

1. .8145, .7491, .7016, .7036, 1.1421,

Log catchability residuals.


## Table 3.3.10.b (Cont'd)

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age , | 5, | 6, | 7, | 8 |
| :---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -7.9087, | -7.8021, | -7.7991, | -7.9100, |
| S.E (Log q), | .2074, | .2433, | .3170, | .3918, |

Regression statistics :
Ages with $q$ dependent on year class strength
Age, Slope, t-value, Intercept, RSquare, No Pts, Reg s.e, Mean Log q

| 0, | .47, | 3.809, | 11.60, | .75, | 19, | .24, | -10.66, |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1, | .54, | 6.766, | 10.61, | .93, | 19, | .13, | -9.20, |
| 2, | .65, | 3.953, | 9.71, | .88, | 19, | .16, | -8.44, |
| 3, | .64, | 3.782, | 9.49, | .87, | 19, | .17, | -8.17, |
| 4, | .69, | 3.450, | 9.11, | .88, | 19, | .17, | -7.98, |

Ages with $q$ independent of year class strength and constant w.r.t. time.
Age, Slope, t-value, Intercept, RSquare, No Pts, Reg s.e, Mean $Q$

| 5, | .91, | .912, | 8.21, | .85, | 19, | .19, | -7.91, |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 6, | .92, | .715, | 8.03, | .82, | 19, | .23, | -7.80, |
| 7, | .94, | .420, | 7.93, | .71, | 19, | .30, | -7.80, |
| 8, | 1.06, | -.307, | 7.85, | .58, | 19, | .43, | -7.91, |

1

Terminal year survivor and $F$ summaries :
Age 0 Catchability dependent on age and year class strength
Year class $=2002$


Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year, | s.e, | s.e, | Ratio, | Ration |  |
| $282388 .$, | .25, | .13, | 2, | .520, | .000 |

Age 1 Catchability dependent on age and year class strength
Year class $=2001$


Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | 年 | Ratio, |  |
| $95403 .$, | .19, | .29, | 3, | 1.518, | .000 |



## Table 3.3.10.b (Cont'd)

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | :--- | :--- | :--- | :--- | :--- |
| at end of year, | s.e, | s.e, | Ratio, |  |  |
| $198028 .$, | .16, | .06, | 4, | .362, | .000 |

Age 3 Catchability dependent on age and year class strength
Year class $=1999$

| Fleet, SMB. 'Tot | , | Estimated, Survivors, 145278., |  |  | Ext, s.e, .116, | Var, <br> Ratio, <br> . 77, | $\begin{gathered} \mathrm{N}, \\ 4^{\prime}, \end{gathered}$ | Scaled, Weights, .818, | $\begin{gathered} \text { Estimated } \\ \text { F } \\ .036 \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| P shrinkage mean | , | 126060 , |  | , , |  |  |  | . 105, | . 042 |
| F shrinkage mean |  | 80935., |  | , , |  |  |  | . 077 , | . 064 |
| Weighted prediction : |  |  |  |  |  |  |  |  |  |
| ```Survivors, at end of year, 136855.,``` | Int, s.e, .14, | Ext, s.e, .11, | $\begin{aligned} & N_{1}^{\prime}, \\ & 6^{\prime}, \end{aligned}$ | $\begin{array}{r} \text { Var, } \\ \text { Ratio, } \\ .812, \end{array}$ | F .038 |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |
| Age 4 Catchability |  | dependent on age and year class strength |  |  |  |  |  |  |  |
| Year class $=1998$ |  |  |  |  |  |  |  |  |  |
| Fleet, |  | $\begin{array}{r} \text { Estimated, } \\ \text { Survivors, } \\ 88149 ., \end{array}$ | $\begin{aligned} & \text { Int, } \\ & \text { s.e, } \\ & .134, \end{aligned}$ |  | Ext, s.e, .019, | Var, <br> Ratio, .14, | N, Scaled, <br> , Weights, <br> 5, .824, |  | $\begin{gathered} \text { Estimated } \\ \text { F } \\ .174 \end{gathered}$ |
| SMB. 'Tot | , |  |  |  |  |  |  |  |  |
| $P$ shrinkage mean | , | 82645. | . 43, , , , |  |  |  | .101, |  | .184 |
| F shrinkage mean | , | 78652., | . 50,1, |  |  |  | .075, . 193 |  |  |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | S.e, | , | Ratio, |  |
| $86829 .$, | .12, | .02, | 7, | .167, | .176 |

Age 5 Catchability constant w.r.t. time and dependent on age
Year class $=1997$


Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| $52048 .$, | .12, | .04, | 7, | .359, | .352 |

1
1 Age 6 Catchability constant w.r.t. time and dependent on age
Year class $=1996$

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | S.e, | , | Ratio, |  |
| $8548 .$, | .12, | .07, | 8, | .604, | .547 |

## Table 3.3.10.b (Cont'd)

Age 7 Catchability constant w.r.t. time and dependent on age
Year class $=1995$

| Fleet, |  | Estimated, Survivors, | Int, s.e, | Ext, s.e, | Var, <br> Ratio, | N, | Scaled, Weights, | $\begin{gathered} \text { Estimated } \\ \mathrm{F} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. Tot | , | 8572., | . 124, | . 075, | . 61, | 8, | . 812, | . 665 |
| $F$ shrinkage mean |  | 6636. | . 50, |  |  |  | . 188, | .797 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| $8169 .$, | .14, | .07, | 9, | .540, | .688 |

1
Age 8 Catchability constant w.r.t. time and dependent on age
Year class $=1994$

| Fleet, |  | Estimated, Survivors, | Int, | Ext, <br> S.e, | Var, <br> Ratio, |  | Scaled, Weights, | $\begin{gathered} \text { Estimated } \\ \mathrm{F} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. Tot | , | 1700., | .139, | . 122, | .88, | 9, | . 721 , | . 782 |
| F shrinkage mean | , | 1491., | . 50, |  |  |  | . 279 , | 855 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | Ratio, | Ration |  |
| $1639 .$, | .17, | .10, | $10^{\prime}$, | .585, | .802 |

Age 9 Catchability constant w.r.t. time and dependent on age
Year class $=1993$

| Fleet, |  | Estimated, Survivors, | Int, |  | Ext, <br> S.e, <br> .094, | Var, <br> Ratio, <br> . 64, | $\begin{aligned} & \mathrm{N}, \\ & 9^{\prime}, \end{aligned}$ | ```Scaled, Weights, .459,``` | $\begin{gathered} \text { Estimated } \\ \text { F } \\ .968 \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. Tot | , | 749., |  |  |  |  |  |  |  |
| F shrinkage mean | , | 869., |  | , , , |  |  |  | . 541 , | . 879 |
| Weighted prediction : |  |  |  |  |  |  |  |  |  |
| Survivors, | Int, | Ext, | N, | Var, | F |  |  |  |  |
| at end of year, | s.e, | s.e, |  | Ratio, | 919 |  |  |  |  |

1
Age 10 Catchability constant w.r.t. time and dependent on age
Year class $=1992$

| Fleet, |  | Estimated, | Int, | Ext, s.e, | Var, <br> Ratio, |  | Scaled, Weights, | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. Tot | , | $305 .,$ | $\begin{aligned} & \text { s.e, } \\ & .128, \end{aligned}$ | $.089$ | $.69$ | $9^{\prime}$ | $.292,$ | $1.112$ |
| F shrinkage mean |  | 399., | . 50, |  |  |  | . 708 , | . 942 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | S.e, | Ratio, | Ration |  |
| $369 .$, | .36, | .09, | 10, | .246, | .990 |

Age 11 Catchability constant w.r.t. time and dependent on age
Year class = 1991


Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | Ratio, |  |  |
| $65 .$, | .42, | .04, | 10, | .106, | 1.000 |

## Table 3.3.10.b (Cont'd)

1
Age 12 Catchability constant w.r.t. time and age (fixed at the value for age) 11
Year class $=1990$

| Fleet, | Estimated, | Int, | Ext, | Var, | N, | Scaled, | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors, | s.e, | s.e, | Ratio, |  | Weights, | F |
| SMB. Tot | 23., | .134, | .067, | . 50, | 9 | .042, | . 940 |
| F shrinkage mean | 19., | . 50, |  |  |  | . 958, | 1.057 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | ry | Ratio, |  |
| $19 .$, | .48, | .06, | 10, | .129, | 1.052 |

Age 13 Catchability constant w.r.t. time and age (fixed at the value for age) 11
Year class $=1989$

| Fleet, |  | Estimated, Survivors, | Int, s.e, | Ext, <br> s.e, | Var, <br> Ratio, | N, | Scaled, Weights, | $\begin{gathered} \text { Estimated } \\ \mathrm{F} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. Tot |  | 24., | . 124, | . 088, | . 71, | 9, | . 041 , | . 939 |
| F shrinkage mean |  | 21., | . 50, |  |  |  | . 959, | 1.012 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | Ratio, | Ration |  |
| $21 .$, | .48, | .04, | $10^{\prime}$, | .087, | 1.009 |

1
Age 14 Catchability constant w.r.t. time and age (fixed at the value for age) 11
Year class $=1988$

| Fleet, |  | Estimated, Survivors, | Int, s.e, | Ext, s.e, | Var, <br> Ratio, | N, | Scaled, Weights, | $\begin{aligned} & \text { Estimated } \\ & \mathrm{F} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SMB. Tot | , | 1., | . 135, | .127, | . 94 , | 9, | .005, | . 000 |
| F shrinkage mean |  | 1., | . 50, |  |  |  | . 995 , | 1.024 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | Ratio, |  |  |
| $1 .$, | .50, | .01, | $10^{\prime}$, | .021, | 1.024 |

Table 3.3.11

## Input data and estimated parameters:

Data: Catch-at-age 1971-2001 and spring trawl survey indices (weighted geometric of North and South) 1985-2003.

## Estimated stock in numbers and total biomass:

| Year/age 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | BIOM(4-11) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984183673. | 76923. | 40551. | 18322. | 13818. | 4980. | 992. | 986. | 879.7 |
| 1985109394. | 121966. | 45023. | 19717. | 7805. | 4732. | 1706. | 341. | 902.6 |
| 1986107960. | 67190. | 66715. | 20919. | 7965. | 2918. | 1685. | 596. | 831.9 |
| 1987246689. | 69095. | 31789. | 26508. | 7064. | 2521. | 913. | 533. | 993.8 |
| 1988228463. | 148296. | 32818. | 11772. | 8274. | 2079. | 748. | 272. | 1041.7 |
| 1989135473. | 147546. | 73365. | 11868. | 3479. | 1853. | 486. | 180. | 1022.6 |
| 199064412. | 84643. | 102641. | 32302. | 4524. | 1185. | 645. | 168. | 821.0 |
| 1991102415. | 41660. | 43613. | 43779. | 11908. | 1580. | 410. | 223. | 686.6 |
| 199277035. | 60898. | 19835. | 15947. | 13997. | 3878. | 502. | 126. | 532.2 |
| 1993133864. | 43115. | 26043. | 6637. | 4481. | 4256. | 1343. | 172. | 575.1 |
| 1994100764. | 79742. | 20832. | 9959. | 2226. | 1242. | 1110. | 351. | 576.6 |
| 199556234. | 61799. | 46324. | 10641. | 4357. | 868. | 485. | 436. | 558.2 |
| 1996121108. | 37566. | 35533. | 25226. | 5229. | 2078. | 411. | 226. | 675.9 |
| 1997132120. | 85831. | 23526. | 18089. | 12222. | 2337. | 933. | 183. | 792.7 |
| 199867608. | 92966. | 54538. | 12997. | 8056. | 4970. | 897. | 356. | 722.0 |
| 1999127193. | 47888. | 53845. | 26722. | 5476. | 2940. | 1589. | 293. | 725.5 |
| 200046939. | 85735. | 25606. | 21927. | 10446. | 2081. | 1125. | 614. | 564.0 |
| 2001147489. | 32255. | 45627. | 10929. | 7713. | 3673. | 747. | 405. | 683.8 |
| 2002132133. | 98772. | 17868. | 19967. | 4093. | 2688. | 1246. | 251. | 737.7 |
| 2003126433. | 91111. | 58164. | 8814. | 8263. | 1498. | 962. | 444. | 799.0 |

## Standard deviation of stock estimate:

| 2002 | 9316. | 5467. | 935. | 1138. | 279. | 247. | 152. |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 11279. | 7572. | 4373. | 746. | 865. | 214. | 180. |

Estimated fishing mortality rates:

| Year/age | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | FBAR(5-10) |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0.209 | 0.336 | 0.520 | 0.653 | 0.871 | 0.870 | 0.863 | 0.871 | 0.685 |
| 1985 | 0.286 | 0.403 | 0.567 | 0.705 | 0.784 | 0.832 | 0.852 | 0.855 | 0.690 |
| 1986 | 0.246 | 0.538 | 0.723 | 0.886 | 0.949 | 0.957 | 0.949 | 0.953 | 0.834 |
| 1987 | 0.306 | 0.543 | 0.785 | 0.962 | 1.022 | 1.015 | 1.009 | 1.015 | 0.889 |
| 1988 | 0.237 | 0.500 | 0.809 | 1.019 | 1.259 | 1.218 | 1.191 | 1.197 | 0.999 |
| 1989 | 0.261 | 0.454 | 0.612 | 0.747 | 0.874 | 0.852 | 0.856 | 0.874 | 0.733 |
| 1990 | 0.236 | 0.454 | 0.651 | 0.798 | 0.851 | 0.862 | 0.863 | 0.869 | 0.746 |
| 1991 | 0.320 | 0.540 | 0.775 | 0.917 | 0.912 | 0.938 | 0.961 | 0.945 | 0.841 |
| 1992 | 0.378 | 0.649 | 0.890 | 1.023 | 0.962 | 0.857 | 0.871 | 0.888 | 0.875 |
| 1993 | 0.318 | 0.526 | 0.760 | 0.893 | 1.041 | 1.092 | 1.105 | 1.099 | 0.903 |
| 1994 | 0.276 | 0.343 | 0.472 | 0.627 | 0.742 | 0.738 | 0.734 | 0.742 | 0.609 |
| 1995 | 0.203 | 0.340 | 0.408 | 0.510 | 0.538 | 0.540 | 0.558 | 0.562 | 0.482 |
| 1996 | 0.143 | 0.268 | 0.447 | 0.522 | 0.595 | 0.597 | 0.606 | 0.606 | 0.506 |
| 1997 | 0.149 | 0.254 | 0.393 | 0.576 | 0.691 | 0.742 | 0.749 | 0.750 | 0.567 |
| 1998 | 0.145 | 0.338 | 0.514 | 0.661 | 0.758 | 0.874 | 0.844 | 0.823 | 0.665 |
| 1999 | 0.194 | 0.426 | 0.680 | 0.737 | 0.761 | 0.740 | 0.731 | 0.738 | 0.679 |
| 2000 | 0.174 | 0.431 | 0.651 | 0.831 | 0.842 | 0.821 | 0.815 | 0.816 | 0.732 |
| 2001 | 0.201 | 0.382 | 0.623 | 0.770 | 0.851 | 0.878 | 0.888 | 0.874 | 0.732 |
| 2002 | 0.172 | 0.330 | 0.504 | 0.680 | 0.799 | 0.821 | 0.825 | 0.821 | 0.660 |

## Standard deviations of $\log (\mathbf{F})$ :

| 2002 | 0.09 | 0.08 | 0.09 | 0.09 | 0.12 | 0.14 | 0.15 | 0.15 | 0.087 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Table 3.3.11 (Continued)
Standardized catch prediction errors:

| Year/age | $\mathbf{4}$ | 5 | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | 9 | 10 | 11 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1985 | 1.47 | 0.91 | 1.04 | -0.64 | 0.60 | -1.16 | 0.85 | 0.75 |
| 1986 | -0.53 | 1.03 | 1.55 | 1.19 | 0.19 | 1.29 | -0.61 | 0.63 |
| 1987 | 1.78 | 1.31 | -1.25 | 0.61 | 0.19 | 0.34 | 0.98 | -0.30 |
| 1988 | -1.72 | 1.07 | 1.94 | -1.14 | 0.68 | 0.23 | 0.24 | 2.05 |
| 1989 | -0.79 | -0.86 | -0.49 | 0.25 | -0.52 | -2.27 | -1.86 | -0.84 |
| 1990 | 0.02 | -1.95 | -0.68 | 1.18 | 1.78 | 0.15 | -0.02 | 0.83 |
| 1991 | 1.29 | 1.56 | -0.38 | 0.47 | -0.27 | 0.54 | 0.44 | 0.23 |
| 1992 | 0.95 | 1.21 | 1.28 | -0.73 | -1.01 | -1.06 | -0.75 | -1.51 |
| 1993 | 0.10 | -1.41 | -0.19 | -0.61 | -1.21 | -0.55 | 1.82 | 1.72 |
| 1994 | 0.39 | -0.58 | -0.67 | 0.36 | 1.75 | -0.94 | -1.04 | 0.81 |
| 1995 | -0.60 | -1.57 | -0.98 | -0.46 | -1.24 | -1.33 | -0.87 | -0.74 |
| 1996 | -0.36 | -0.89 | -1.16 | 0.36 | 0.08 | 0.27 | 1.40 | -0.18 |
| 1997 | -0.87 | 0.63 | 0.00 | -1.17 | 0.88 | 0.17 | 0.91 | 1.61 |
| 1998 | -0.21 | 0.97 | 2.02 | 1.09 | -1.46 | 0.76 | 0.00 | 0.04 |
| 1999 | 1.94 | 1.85 | 0.35 | 1.06 | 0.30 | -1.60 | -1.90 | -1.58 |
| 2000 | -0.09 | 1.08 | 0.67 | -0.88 | 0.66 | 0.08 | -0.68 | -0.49 |
| 2001 | 0.36 | 0.11 | -0.25 | 0.52 | -1.08 | 0.11 | 0.49 | -0.46 |
| 2002 | -0.77 | -1.53 | 0.24 | -0.18 | 1.29 | 0.10 | 0.46 | -0.10 |

Skewness and kurtosis (Standardized normal distribution): 0.817-1.904
Correlation between cohorts aages and years: $\begin{array}{llll}0.28 & 0.31 & -0.10\end{array}$
Standardized prediction errors of cpue:

| Year/age | $\mathbf{4}$ | 5 | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 1985 | 2.01 | 0.30 | 0.85 | 0.64 | 1.34 | 0.8 |
| 1986 | -0.99 | -0.42 | -0.52 | -0.66 | -1.59 | -0.35 |
| 1987 | -0.08 | -0.65 | -0.96 | -0.19 | 0.28 | -0.02 |
| 1988 | 0.67 | 1.49 | -0.85 | -0.46 | 0.58 | -0.29 |
| 1989 | 1.38 | 0.94 | 1.91 | 1.12 | 0.39 | -0.11 |
| 1990 | -1.03 | -1.50 | -0.65 | 1.65 | 0.94 | 0.88 |
| 1991 | 0.54 | 1.70 | -0.47 | 1.29 | 0.16 | 1.09 |
| 1992 | 0.86 | -0.72 | -0.72 | -1.12 | -0.37 | -0.25 |
| 1993 | 0.73 | 0.54 | 0.41 | -0.97 | -0.67 | -1.09 |
| 1994 | -0.75 | -0.68 | -0.89 | -1.08 | -0.78 | -0.56 |
| 1995 | -0.44 | -0.54 | 0.21 | -0.18 | 0.01 | 0.30 |
| 1996 | 0.75 | 1.04 | -0.96 | 1.76 | 2.23 | 1.21 |
| 1997 | 0.11 | 0.14 | -0.17 | -1.32 | -0.08 | -1.68 |
| 1998 | -0.24 | 0.06 | 0.10 | -0.15 | -0.83 | 0.43 |
| 1999 | 0.87 | -0.36 | -0.79 | 0.18 | 0.40 | -1.25 |
| 2000 | 0.10 | 0.39 | -0.24 | -1.46 | 0.24 | -1.17 |
| 2001 | -0.38 | -0.84 | -0.81 | -1.03 | -1.99 | -1.15 |
| 2002 | 0.44 | 0.93 | 2.77 | 0.53 | 1.05 | -0.35 |
| 2003 | -0.32 | -0.14 | -0.78 | 2.29 | -0.02 | 1.28 |

Skewness and kurtosis (Standardized normal distribution): 0.011 -0.473
Correlation within cohorts: 0.10
Correlation between ages and years: 0.35 -0.14

Table 3.3.12 AD Model Builder -Statistical Catch-at-age Model- AD-CAM - diagnostic and results.

## Input data and estimated parameters:

- The model used catchdata from 1955 to 2002 and survey data from 1985-2003. Age groups included are 1-10 in the survey and 3-14 in the catches.

Parameter settings and assumptions used:

- Fishing mortality was estimated for every year and age.
- Recruitment was assumed to be lognormally distributed around a Ricker curve with the CV of the lognormal distribution estimated. Timetrend in $\mathrm{R}_{\max }$ of the Ricker curve was allowed and CV of the residuals in the SSBrecruitment relationship depend on stock size. The SSB - recruitment relationship was based on spawning stock based on maturity-at-age from the survey, predicting the survey maturitity at age backwards in time from the observations from the catches.
- Migrations for specified years in specified ages are estimated.
- Catchability in the survey was dependent on stocksize for ages 1-5.
- CV of commercial catch data and of survey indices as function of age are estimated. The CV of the commercial catch is a parabola but estimated seperately for each age in the survey (change from last year when it was also a $2^{\text {nd }}$ order polynomial) Correlation of residuals of different agegroups in the survey was estimated as a $1^{\text {st }}$ order AR model.
- Fishing mortality of each age group was random walk with standard deviation specified as proportion of the estimated CV in the catch-at-age data. In the input file the process error (variablility in F) is specified to be larger than the measurement error for the younger ages but the measurement error is specified to be larger for the older age groups.
- The model estimates standard deviation on survey and age disaggregated catches. The division of the standard deviation in catches between process (random walk of F ) and measurement error must be specified.

Some non-traditional of the assemssment model are.

- Rmax decrease by $0.9 \%$ per year from 1955 to 1995 so precdicted recruitment in 1995 is expected to be $67 \%$ of what it was in 1995 for the same spawing size of the spawing stock. At lesat part of this trend is considered to be due to different composition of the spawning stock with higher percentage of young fish in the spawing stock in recent years. Using catch maturity-at-age gives $1.5 \%$ trend per year.
- CV in recruitment. increases with reduced spawning stock as expected.
- Spawning stock giving maximum recruitment at about 400 kT . (500-550 kT using catch maturity-at-age)

| Age | $\mathbf{M}$ | Survey <br> sigma | Survey <br> lnQ | Survey <br> Power | Meansel | Progsel | Sigma |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1}$ | 0.2 | 0.364 | -22.716 | 2.028 | -1.000 | -1.000 | -1.000 |
| $\mathbf{2}$ | 0.2 | 0.138 | -18.866 | 1.839 | -1.000 | -1.000 | -1.000 |
| $\mathbf{3}$ | 0.2 | 0.183 | -15.571 | 1.607 | 0.148 | 0.033 | 0.182 |
| $\mathbf{4}$ | 0.2 | 0.198 | -13.552 | 1.476 | 0.502 | 0.117 | 0.145 |
| $\mathbf{5}$ | 0.2 | 0.167 | -11.499 | 1.331 | 0.773 | 0.249 | 0.122 |
| $\mathbf{6}$ | 0.2 | 0.139 | -7.703 | 1.000 | 0.953 | 0.397 | 0.110 |
| $\mathbf{7}$ | 0.2 | 0.168 | -7.548 | 1.000 | 1.185 | 0.481 | 0.105 |
| $\mathbf{8}$ | 0.2 | 0.207 | -7.516 | 1.000 | 1.400 | 0.531 | 0.106 |
| $\mathbf{9}$ | 0.2 | 0.238 | -7.650 | 1.000 | 1.473 | 0.561 | 0.114 |
| $\mathbf{1 0}$ | 0.2 | 0.234 | -7.623 | 1.000 | 1.573 | 0.597 | 0.130 |
| $\mathbf{1 1}$ | 0.2 | -1.000 | -1.000 | -1.000 | 1.520 | 0.584 | 0.157 |
| $\mathbf{1 2}$ | 0.2 | -1.000 | -1.000 | -1.000 | 1.355 | 0.612 | 0.202 |
| $\mathbf{1 3}$ | 0.2 | -1.000 | -1.000 | -1.000 | 1.000 | 1.000 | 0.276 |
| $\mathbf{1 4}$ | 0.2 | -1.000 | -1.000 | -1.000 | 1.000 | 1.000 | 0.401 |

Table 3.3.12 (Continued)
Estimated fishing mortality rates:

| Year/age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | 0.086 | 0.088 | 0.149 | 0.101 | 0.067 | 0.037 | 0.026 | 0.029 | 0.051 | 0.066 | 0.069 | 0.044 |
| $\mathbf{4}$ | 0.318 | 0.387 | 0.312 | 0.294 | 0.185 | 0.147 | 0.146 | 0.146 | 0.186 | 0.193 | 0.218 | 0.177 |
| $\mathbf{5}$ | 0.548 | 0.590 | 0.504 | 0.323 | 0.359 | 0.274 | 0.265 | 0.367 | 0.406 | 0.439 | 0.400 | 0.357 |
| $\mathbf{6}$ | 0.797 | 0.843 | 0.734 | 0.561 | 0.394 | 0.454 | 0.445 | 0.526 | 0.696 | 0.657 | 0.645 | 0.610 |
| $\mathbf{7}$ | 0.906 | 1.022 | 0.865 | 0.732 | 0.525 | 0.531 | 0.614 | 0.664 | 0.724 | 0.870 | 0.792 | 0.749 |
| $\mathbf{8}$ | 0.884 | 1.002 | 1.072 | 0.927 | 0.566 | 0.614 | 0.680 | 0.765 | 0.774 | 0.860 | 0.915 | 0.877 |
| $\mathbf{9}$ | 0.815 | 0.817 | 1.018 | 0.953 | 0.575 | 0.626 | 0.725 | 0.834 | 0.844 | 0.868 | 0.927 | 0.952 |
| $\mathbf{1 0}$ | 0.781 | 0.789 | 0.853 | 0.854 | 0.785 | 0.800 | 0.838 | 0.874 | 0.897 | 0.943 | 0.986 | 1.007 |
| $\mathbf{1 1}$ | 0.696 | 0.702 | 0.720 | 0.720 | 0.857 | 0.871 | 0.891 | 0.904 | 0.905 | 0.919 | 0.936 | 0.939 |
| $\mathbf{1 2}$ | 0.655 | 0.653 | 0.655 | 0.657 | 0.955 | 0.952 | 0.955 | 0.964 | 0.966 | 0.971 | 0.966 | 0.957 |
| $\mathbf{1 3}$ | 0.463 | 0.463 | 0.463 | 0.463 | 1.578 | 1.578 | 1.578 | 1.578 | 1.576 | 1.574 | 1.572 | 1.572 |
| $\mathbf{1 4}$ | 0.463 | 0.463 | 0.463 | 0.463 | 1.578 | 1.578 | 1.578 | 1.578 | 1.576 | 1.574 | 1.572 | 1.572 |
| $\mathbf{F}(\mathbf{5 - 1 0})$ | $\mathbf{0 . 7 8 8}$ | $\mathbf{0 . 8 4 4}$ | $\mathbf{0 . 8 4 1}$ | $\mathbf{0 . 7 2 5}$ | $\mathbf{0 . 5 3 4}$ | $\mathbf{0 . 5 5 0}$ | $\mathbf{0 . 5 9 4}$ | $\mathbf{0 . 6 7 2}$ | $\mathbf{0 . 7 2 3}$ | $\mathbf{0 . 7 7 3}$ | $\mathbf{0 . 7 7 8}$ | $\mathbf{0 . 7 5 9}$ |

Estimated stock in numbers (millions):

| Year/age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 215.680 | 110.887 | 221.084 | 270.184 | 121.910 | 232.422 | 86.610 | 274.916 | 247.375 | 248.890 | 308.868 | 103.404 | 291.359 |
| 2 | 202.306 | 176.583 | 90.787 | 181.008 | 221.208 | 99.812 | 190.291 | 70.911 | 225.082 | 202.534 | 203.774 | 252.880 | 84.660 |
| 3 | 107.719 | 165.634 | 144.574 | 74.330 | 148.197 | 181.110 | 81.719 | 155.797 | 58.057 | 184.281 | 165.821 | 166.836 | 207.040 |
| 4 | 100.148 | 80.933 | 124.133 | 101.950 | 55.008 | 113.487 | 142.924 | 65.170 | 123.969 | 45.182 | 141.272 | 126.674 | 130.668 |
| 5 | 43.325 | 59.642 | 44.983 | 74.422 | 62.216 | 37.425 | 80.194 | 101.122 | 46.128 | 84.247 | 30.493 | 93.043 | 86.862 |
| 6 | 44.697 | 20.516 | 27.066 | 22.238 | 44.107 | 35.561 | 23.306 | 50.378 | 57.369 | 25.170 | 44.456 | 16.738 | 53.290 |
| 7 | 38.486 | 16.495 | 7.232 | 10.637 | 10.386 | 24.352 | 18.488 | 12.230 | 24.386 | 23.424 | 10.681 | 19.097 | 7.449 |
| 8 | 12.310 | 12.739 | 4.862 | 2.493 | 4.190 | 5.028 | 11.719 | 8.193 | 5.156 | 9.683 | 8.037 | 3.960 | 7.390 |
| 9 | 1.554 | 4.165 | 3.831 | 1.363 | 0.807 | 1.948 | 2.229 | 4.862 | 3.120 | 1.947 | 3.353 | 2.635 | 1.349 |
| 10 | 0.455 | 0.563 | 1.507 | 1.133 | 0.430 | 0.372 | 0.853 | 0.884 | 1.728 | 1.098 | 0.669 | 1.086 | 0.833 |
| 11 | 0.304 | 0.171 | 0.210 | 0.526 | 0.395 | 0.161 | 0.137 | 0.302 | 0.302 | 0.577 | 0.350 | 0.204 | 0.325 |
| 12 | 0.098 | 0.124 | 0.069 | 0.084 | 0.209 | 0.137 | 0.055 | 0.046 | 0.100 | 0.100 | 0.188 | 0.112 | 0.065 |
| 13 | 0.056 | 0.042 | 0.053 | 0.029 | 0.035 | 0.066 | 0.043 | 0.017 | 0.014 | 0.031 | 0.031 | 0.059 | 0.035 |
| 14 | 0.039 | 0.029 | 0.021 | 0.027 | 0.015 | 0.006 | 0.011 | 0.007 | 0.003 | 0.002 | 0.005 | 0.005 | 0.010 |

Table 3.3.12 (Continued)

## Residuals:

## Log(Cay-observed/Cay-predicted)

| Year/age | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | 0.00 | 0.08 | -0.06 | -0.02 | -0.12 | -0.03 | 0.06 | -0.04 | 0.12 | -0.05 | 0.21 |
| $\mathbf{4}$ | 0.04 | -0.04 | 0.03 | -0.04 | -0.10 | -0.11 | 0.01 | -0.09 | 0.09 | 0.02 | 0.07 |
| $\mathbf{5}$ | -0.02 | -0.01 | 0.05 | -0.03 | 0.10 | -0.14 | -0.08 | 0.09 | -0.07 | 0.05 | -0.02 |
| $\mathbf{6}$ | 0.01 | 0.01 | -0.05 | 0.18 | -0.11 | 0.08 | -0.05 | 0.06 | 0.03 | -0.23 | 0.00 |
| $\mathbf{7}$ | -0.08 | 0.10 | 0.03 | -0.06 | -0.04 | -0.05 | 0.17 | 0.03 | -0.07 | -0.14 | 0.06 |
| $\mathbf{8}$ | -0.06 | 0.01 | 0.02 | 0.05 | -0.17 | 0.06 | -0.04 | 0.11 | -0.06 | -0.10 | -0.04 |
| $\mathbf{9}$ | -0.11 | 0.09 | 0.03 | 0.18 | -0.30 | -0.05 | 0.13 | -0.06 | 0.18 | -0.25 | -0.03 |
| $\mathbf{1 0}$ | 0.03 | -0.09 | 0.10 | 0.12 | -0.25 | -0.17 | 0.07 | -0.09 | 0.40 | -0.16 | -0.15 |
| $\mathbf{1 1}$ | -0.03 | 0.08 | -0.22 | 0.44 | -0.29 | 0.09 | -0.11 | -0.54 | 0.60 | 0.13 | -0.29 |
| $\mathbf{1 2}$ | -0.25 | 0.08 | -0.07 | 0.17 | 0.09 | -0.14 | 0.38 | -0.53 | 0.11 | 0.28 | 0.17 |
| $\mathbf{1 3}$ | 0.31 | -0.31 | 0.33 | 0.78 | -0.15 | -0.55 | -0.54 | -0.44 | -0.49 | 0.59 | 0.10 |
| $\mathbf{1 4}$ | 0.35 | 0.13 | 0.08 | 0.39 | -1.68 | -0.12 | -0.10 | -0.20 | 0.59 | -0.43 | -0.34 |


| Year/age | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | -0.10 | -0.11 | -0.11 | 0.05 | 0.00 | 0.12 | -0.10 |
| $\mathbf{4}$ | 0.05 | -0.07 | -0.04 | 0.10 | -0.07 | 0.00 | -0.01 |
| $\mathbf{5}$ | -0.10 | 0.02 | -0.11 | 0.07 | 0.08 | 0.02 | -0.05 |
| $\mathbf{6}$ | 0.04 | -0.13 | 0.07 | -0.03 | 0.03 | 0.01 | -0.02 |
| $\mathbf{7}$ | 0.03 | -0.05 | 0.08 | 0.06 | -0.08 | 0.05 | -0.03 |
| $\mathbf{8}$ | 0.02 | 0.12 | -0.04 | -0.01 | 0.12 | -0.12 | 0.05 |
| $\mathbf{9}$ | 0.01 | 0.08 | 0.16 | -0.16 | 0.08 | 0.08 | -0.10 |
| $\mathbf{1 0}$ | 0.10 | 0.10 | 0.05 | -0.08 | -0.07 | 0.15 | 0.08 |
| $\mathbf{1 1}$ | -0.12 | 0.53 | 0.17 | -0.16 | 0.00 | 0.03 | 0.07 |
| $\mathbf{1 2}$ | -0.18 | -0.11 | 0.74 | -0.13 | 0.27 | 0.09 | -0.47 |
| $\mathbf{1 3}$ | 0.12 | 0.15 | 0.95 | -0.06 | 0.00 | -0.56 | -0.05 |
| $\mathbf{1 4}$ | 0.13 | 0.94 | 0.88 | 0.84 | 1.21 | 0.84 | -0.66 |

Table 3.3.12 (Continued)
Log(Uay-observed/Uay-predicted)

| Year/age | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | -0.39 | 0.55 | 0.63 | -0.24 | 0.29 | -0.27 | -0.33 | -0.68 | -0.48 | 0.50 | -0.38 |
| $\mathbf{2}$ | 0.21 | -0.15 | 0.04 | 0.04 | 0.10 | 0.10 | -0.37 | -0.07 | -0.11 | -0.24 | 0.05 |
| $\mathbf{3}$ | 0.20 | -0.27 | 0.01 | 0.46 | 0.47 | 0.01 | -0.07 | -0.15 | 0.01 | -0.16 | -0.27 |
| $\mathbf{4}$ | 0.43 | -0.28 | -0.32 | 0.03 | 0.53 | -0.09 | 0.12 | -0.02 | -0.01 | -0.04 | -0.25 |
| $\mathbf{5}$ | 0.25 | -0.07 | -0.05 | -0.07 | -0.05 | -0.19 | 0.22 | -0.13 | 0.01 | -0.20 | 0.13 |
| $\mathbf{6}$ | 0.30 | 0.05 | 0.02 | -0.43 | 0.11 | -0.14 | 0.09 | -0.14 | -0.02 | -0.40 | 0.00 |
| $\mathbf{7}$ | 0.53 | -0.23 | 0.04 | 0.13 | -0.24 | -0.04 | 0.25 | -0.18 | -0.37 | -0.25 | -0.23 |
| $\mathbf{8}$ | 0.23 | -0.32 | -0.12 | 0.52 | -0.17 | -0.28 | -0.24 | -0.06 | -0.16 | -0.44 | -0.11 |
| $\mathbf{9}$ | 0.50 | -0.37 | -0.07 | -0.24 | 0.27 | 0.09 | 0.28 | -0.23 | -0.20 | -0.07 | 0.06 |
| $\mathbf{1 0}$ | 0.56 | 0.00 | -0.12 | -0.29 | -0.02 | 0.28 | 0.56 | -0.03 | -0.56 | -0.04 | -0.02 |


| Year/age | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | -0.54 | 0.33 | -0.11 | 0.01 | 0.94 | 0.06 | -0.30 | 0.10 |
| $\mathbf{2}$ | -0.16 | 0.06 | 0.48 | 0.14 | 0.17 | -0.10 | 0.07 | -0.07 |
| $\mathbf{3}$ | -0.06 | 0.07 | -0.17 | -0.03 | 0.17 | -0.06 | 0.04 | -0.19 |
| $\mathbf{4}$ | -0.14 | 0.16 | 0.07 | 0.10 | -0.22 | -0.20 | 0.01 | -0.08 |
| $\mathbf{5}$ | 0.17 | -0.03 | 0.44 | -0.06 | -0.02 | -0.49 | 0.01 | -0.05 |
| $\mathbf{6}$ | 0.04 | 0.03 | 0.37 | 0.09 | -0.15 | -0.11 | 0.10 | -0.10 |
| $\mathbf{7}$ | 0.23 | 0.05 | 0.18 | 0.04 | -0.20 | -0.33 | 0.00 | 0.20 |
| $\mathbf{8}$ | 0.49 | 0.21 | 0.35 | 0.02 | -0.03 | -0.58 | -0.16 | 0.32 |
| $\mathbf{9}$ | 0.29 | -0.45 | 0.48 | 0.09 | -0.38 | -0.48 | -0.33 | 0.69 |
| $\mathbf{1 0}$ | 0.24 | 0.06 | 0.64 | -0.03 | -0.35 | -0.12 | -0.33 | -0.39 |

Table 3.3.12 (Continued)
Summary:

| Year | F5-10 | SSB | Bio4+ | N3 | HarvestRatio | Landings |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 0.31 | 1166 | 2323 | 154700 | 0.40 | 545 |
| 1956 | 0.30 | 1094 | 2071 | 182232 | 0.39 | 487 |
| 1957 | 0.31 | 1102 | 1934 | 170017 | 0.38 | 455 |
| 1958 | 0.31 | 1292 | 1996 | 218965 | 0.40 | 517 |
| 1959 | 0.33 | 1071 | 1812 | 301029 | 0.42 | 459 |
| 1960 | 0.35 | 912 | 1792 | 152184 | 0.45 | 470 |
| 1961 | 0.34 | 749 | 1422 | 196010 | 0.43 | 332 |
| 1962 | 0.38 | 789 | 1511 | 132986 | 0.46 | 389 |
| 1963 | 0.44 | 673 | 1286 | 172384 | 0.53 | 409 |
| 1964 | 0.52 | 594 | 1203 | 276699 | 0.60 | 437 |
| 1965 | 0.56 | 461 | 1130 | 247052 | 0.63 | 387 |
| 1966 | 0.55 | 425 | 1211 | 269545 | 0.57 | 353 |
| 1967 | 0.56 | 501 | 1382 | 312476 | 0.50 | 336 |
| 1968 | 0.58 | 582 | 1503 | 170555 | 0.50 | 382 |
| 1969 | 0.54 | 669 | 1489 | 253916 | 0.52 | 403 |
| 1970 | 0.57 | 662 | 1418 | 185905 | 0.61 | 475 |
| 1971 | 0.61 | 495 | 1160 | 185795 | 0.73 | 447 |
| 1972 | 0.66 | 421 | 1008 | 139494 | 0.76 | 391 |
| 1973 | 0.69 | 431 | 830 | 282957 | 0.87 | 369 |
| 1974 | 0.74 | 330 | 916 | 177179 | 0.91 | 368 |
| 1975 | 0.76 | 335 | 883 | 260858 | 0.89 | 365 |
| 1976 | 0.72 | 287 | 949 | 391593 | 0.76 | 346 |
| 1977 | 0.65 | 329 | 1327 | 139528 | 0.61 | 340 |
| 1978 | 0.54 | 394 | 1331 | 224158 | 0.52 | 330 |
| 1979 | 0.49 | 507 | 1421 | 243663 | 0.53 | 366 |
| 1980 | 0.52 | 541 | 1479 | 146789 | 0.58 | 432 |
| 1981 | 0.61 | 428 | 1306 | 144442 | 0.66 | 465 |
| 1982 | 0.71 | 263 | 965 | 133815 | 0.76 | 380 |
| 1983 | 0.72 | 221 | 805 | 224560 | 0.73 | 298 |
| 1984 | 0.66 | 229 | 917 | 140922 | 0.70 | 282 |
| 1985 | 0.68 | 267 | 927 | 136228 | 0.77 | 323 |
| 1986 | 0.77 | 266 | 840 | 342946 | 0.89 | 365 |
| 1987 | 0.83 | 255 | 1038 | 301229 | 0.90 | 390 |
| 1988 | 0.85 | 201 | 1101 | 181450 | 0.83 | 378 |
| 1989 | 0.73 | 286 | 1104 | 86342 | 0.81 | 363 |
| 1990 | 0.73 | 336 | 837 | 128886 | 0.88 | 335 |
| 1991 | 0.79 | 218 | 676 | 107725 | 1.00 | 308 |
| 1992 | 0.84 | 236 | 539 | 165646 | 1.05 | 265 |
| 1993 | 0.84 | 220 | 572 | 144582 | 0.99 | 251 |
| 1994 | 0.73 | 256 | 581 | 74331 | 0.76 | 178 |
| 1995 | 0.53 | 324 | 548 | 148206 | 0.61 | 169 |
| 1996 | 0.55 | 270 | 656 | 181142 | 0.55 | 181 |
| 1997 | 0.59 | 353 | 794 | 81738 | 0.56 | 203 |
| 1998 | 0.67 | 294 | 720 | 155886 | 0.69 | 244 |
| 1999 | 0.72 | 323 | 730 | 58125 | 0.81 | 265 |
| 2000 | 0.77 | 242 | 559 | 184586 | 0.85 | 239 |
| 2001 | 0.78 | 325 | 663 | 166110 | 0.79 | 234 |
| 2002 | 0.76 | 357 | 704 | 167018 | 0.69 | 208 |
| 2003 | 0.57 | 376 | 766 | 207175 | 0.57 |  |
| Mean | 0.61 | 477 | 1125 | 188811 | 0.67 | 352 |

Table 3.3.13

## Input data and estimated parameters:

The fitting procedure was based on minimizing
$S S R_{C}=\sum_{a y} \frac{\left[\ln \left(C_{a y}\right)-\ln \left(\hat{C}_{a y}\right)\right]^{2}}{2 \sigma_{a}^{2}}$
$S S R_{I}=\sum_{a y} \frac{\left[\ln \left(I_{a y}\right)-\ln \left(\hat{I}_{a y}\right)\right]^{2}}{2 \rho_{a}^{2}}$
$S S R_{T}=a_{C} S S R_{c}+a_{I} S S R_{I}$

The weighting factor ( $\mathrm{a}_{\mathrm{c}}$ and $\mathrm{a}_{1}$ ) in each case was set to one. To reflect different accuracy by age in the estimation of catch-at-age and survey-index-at-age the residuals by age in the SSRC and SSRI were weighted according:

| Age -> | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Catch (sigma) |  |  | 0,32 | 0,22 | 0,16 | 0,13 | 0,12 | 0,11 | 0,12 | 0,14 | 0,17 | 0,24 | 0,36 | 0,60 | 0,60 |
| Survey (rho) | 0,39 | 0,30 | 0,24 | 0,22 | 0,20 | 0,21 | 0,22 | 0,26 | 0,32 | 0,43 |  |  |  |  |  |

In order to force the model to fit the observed yield within each year an additional minimization factor (penalty factor) was set in:
$S S R_{Y}=\sum_{y} \frac{\left[\ln \left(Y_{y}\right)-\ln \left(\hat{Y}_{y}\right)\right]^{2}}{2 \sigma_{a}^{2}}$
$S S R_{T}=a_{C} S S R_{c}+a_{I} S S R_{I}+a_{Y} S S R_{y}$
where the ay factor was set to 1000 and the weight for the survey and catch matrices were set to unity.
The input data that went into the model where:

1) catch-in-number matrix was based on ages 3 to 14 for years 1985 to 2002, Ca for ages 1 and 2 were assuemd 0 (thus 216 input values).
2) aged survey indices for ages 1 to 9 for years 1985-2003 (171 input values).
3) auxillary data on the CV by age groups for Cay and Uay - data obtained from hoski@hafro.is (see intext table above).
4) auxillary data such as corresponding weight-at-age in the catch and in the survey and maturity-at-age in the catch.

The number of estimated parameters were:

1) 13 estimates of Na 85 for $\mathrm{a}=2,3, \ldots, 14$
2) 19 estimates of N1y for $\mathrm{y}=1985,1986, \ldots ., 2003$
3) 9 estimates of $\alpha_{a}$ for $a=1,2, \ldots, 9$
4) 7 estimates of $\beta_{\mathrm{a}}$ for $\mathrm{a}=1,2, \ldots$, 7. I.e. power model was used for age groups $1-7$, simple linear model for age groups 8 and 9 was used $(\beta=1)$. The judgement on where to use power model was based on residual plots:
5) 18 estimates of Fy for $\mathrm{y}=1985,1986, \ldots, 2002$
6) estmates of aso and a95
or a total of 68 parameter estimates.

Table 3.3.13 (Continued)
Estimated fishing mortality rates:

| Year/age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ |
| :---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{3}$ | 0.079 | 0.087 | 0.087 | 0.061 | 0.050 | 0.047 | 0.047 | 0.055 | 0.065 | 0.068 | 0.067 |
| $\mathbf{4}$ | 0.245 | 0.267 | 0.269 | 0.188 | 0.155 | 0.145 | 0.145 | 0.170 | 0.201 | 0.209 | 0.207 |
| $\mathbf{5}$ | 0.540 | 0.589 | 0.594 | 0.415 | 0.342 | 0.319 | 0.319 | 0.376 | 0.443 | 0.461 | 0.457 |
| $\mathbf{6}$ | 0.789 | 0.860 | 0.867 | 0.606 | 0.499 | 0.466 | 0.467 | 0.549 | 0.648 | 0.674 | 0.668 |
| $\mathbf{7}$ | 0.898 | 0.979 | 0.987 | 0.689 | 0.568 | 0.531 | 0.531 | 0.624 | 0.737 | 0.767 | 0.760 |
| $\mathbf{8}$ | 0.931 | 1.015 | 1.023 | 0.715 | 0.589 | 0.550 | 0.550 | 0.648 | 0.764 | 0.795 | 0.788 |
| $\mathbf{9}$ | 0.940 | 1.025 | 1.033 | 0.722 | 0.595 | 0.556 | 0.556 | 0.654 | 0.772 | 0.803 | 0.796 |
| $\mathbf{1 0}$ | 0.943 | 1.028 | 1.036 | 0.724 | 0.597 | 0.557 | 0.557 | 0.656 | 0.774 | 0.805 | 0.798 |
| $\mathbf{1 1}$ | 0.944 | 1.028 | 1.037 | 0.725 | 0.597 | 0.558 | 0.558 | 0.656 | 0.774 | 0.806 | 0.799 |
| $\mathbf{1 2}$ | 0.944 | 1.029 | 1.037 | 0.725 | 0.597 | 0.558 | 0.558 | 0.656 | 0.774 | 0.806 | 0.799 |
| $\mathbf{1 3}$ | 0.944 | 1.029 | 1.037 | 0.725 | 0.597 | 0.558 | 0.558 | 0.656 | 0.774 | 0.806 | 0.799 |
| $\mathbf{1 4}$ | 0.944 | 1.029 | 1.037 | 0.725 | 0.597 | 0.558 | 0.558 | 0.656 | 0.774 | 0.806 | 0.799 |
| $\mathbf{F ( 5 - 1 0 )}$ | $\mathbf{0 . 8 4 0}$ | $\mathbf{0 . 9 1 6}$ | $\mathbf{0 . 9 2 3}$ | $\mathbf{0 . 6 4 5}$ | $\mathbf{0 . 5 3 2}$ | $\mathbf{0 . 4 9 7}$ | $\mathbf{0 . 4 9 7}$ | $\mathbf{0 . 5 8 4}$ | $\mathbf{0 . 6 9 0}$ | $\mathbf{0 . 7 1 7}$ | $\mathbf{0 . 7 1 1}$ |

Estimated stock in numbers:

| Year/age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | 184.87 | 113.14 | 230.91 | 261.69 | 135.56 | 251.88 | 95.98 | 281.38 | 258.25 | 265.34 | 305.91 | 100.84 |
| $\mathbf{2}$ | 215.807 | 150.465 | 92.040 | 187.829 | 213.286 | 110.574 | 205.503 | 78.309 | 229.430 | 210.416 | 216.145 | 249.210 |
| $\mathbf{3}$ | 106.331 | 172.814 | 120.249 | 73.542 | 151.186 | 172.192 | 89.352 | 166.062 | 63.133 | 184.455 | 169.045 | 173.674 |
| 200.922 |  |  |  |  |  |  |  |  |  |  |  |  |
| $\mathbf{4}$ | 103.404 | 80.412 | 129.759 | 90.227 | 56.650 | 117.716 | 134.517 | 69.802 | 128.657 | 48.429 | 141.121 | 129.405 |
| $\mathbf{5}$ | 47.819 | 66.260 | 50.402 | 81.157 | 61.199 | 39.719 | 83.385 | 95.284 | 48.195 | 86.146 | 32.164 | 93.892 |
| $\mathbf{6}$ | 44.700 | 22.808 | 30.102 | 22.789 | 43.880 | 35.597 | 23.631 | 49.607 | 53.578 | 25.326 | 44.465 | 16.667 |
| $\mathbf{7}$ | 43.094 | 16.619 | 7.898 | 10.352 | 10.177 | 21.802 | 18.279 | 12.134 | 23.458 | 22.950 | 10.568 | 18.660 |
| $\mathbf{7}$ | 12.855 | 14.377 | 5.114 | 2.411 | 4.253 | 4.721 | 10.501 | 8.804 | 5.321 | 9.193 | 8.730 | 4.046 |
| $\mathbf{8}$ | 1.818 | 4.147 | 4.265 | 1.505 | 0.966 | 1.932 | 2.229 | 4.958 | 3.772 | 2.029 | 3.398 | 3.249 |
| $\mathbf{9}$ | 0.550 | 0.581 | 1.218 | 1.242 | 0.598 | 0.436 | 0.907 | 1.047 | 2.111 | 1.427 | 0.744 | 1.255 |
| $\mathbf{1 0}$ | 0.343 | 0.175 | 0.170 | 0.354 | 0.493 | 0.270 | 0.204 | 0.426 | 0.445 | 0.797 | 0.522 | 0.274 |
| $\mathbf{1 1}$ | 0.092 | 0.109 | 0.051 | 0.049 | 0.140 | 0.222 | 0.126 | 0.096 | 0.181 | 0.168 | 0.292 | 0.192 |
| $\mathbf{1 2}$ | 0.035 | 0.029 | 0.032 | 0.015 | 0.020 | 0.063 | 0.104 | 0.059 | 0.041 | 0.068 | 0.061 | 0.107 |
| $\mathbf{1 3}$ | 0.016 | 0.011 | 0.009 | 0.009 | 0.006 | 0.009 | 0.030 | 0.049 | 0.025 | 0.015 | 0.025 | 0.023 |
| $\mathbf{1 4}$ |  |  |  |  |  |  |  |  |  | 0.046 |  |  |

Table 3.3.13. (Continued)

## Residuals:

Log(Uay-observed/Uay-predicted)

| Year\Age | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | -0.27 | 0.64 | 0.55 | -0.32 | 0.37 | -0.39 | 0.06 | -0.55 | -0.53 | 0.59 | -0.46 |
| $\mathbf{2}$ | 0.17 | 0.02 | 0.13 | -0.12 | 0.01 | 0.12 | -0.47 | 0.25 | -0.09 | -0.29 | 0.13 |
| $\mathbf{3}$ | 0.10 | -0.26 | 0.21 | 0.58 | 0.35 | -0.06 | -0.06 | -0.22 | 0.29 | -0.16 | -0.30 |
| $\mathbf{4}$ | 0.56 | -0.37 | -0.43 | 0.19 | 0.65 | -0.11 | 0.06 | -0.01 | -0.12 | 0.15 | -0.18 |
| $\mathbf{5}$ | 0.21 | -0.03 | -0.14 | -0.26 | 0.12 | -0.15 | 0.19 | -0.25 | -0.03 | -0.27 | 0.25 |
| $\mathbf{6}$ | 0.27 | 0.05 | 0.02 | -0.45 | 0.04 | -0.29 | 0.06 | -0.18 | -0.08 | -0.31 | 0.06 |
| $\mathbf{7}$ | 0.53 | -0.28 | 0.07 | 0.22 | -0.21 | -0.14 | -0.07 | -0.23 | -0.32 | -0.12 | -0.05 |
| $\mathbf{8}$ | 0.22 | -0.29 | -0.11 | 0.64 | -0.15 | -0.30 | -0.25 | -0.17 | -0.21 | -0.38 | -0.02 |
| $\mathbf{9}$ | 0.68 | -0.42 | -0.03 | -0.24 | 0.25 | 0.02 | 0.22 | -0.13 | -0.25 | -0.10 | 0.02 |


| Year $\backslash$ Age | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | -0.68 | 0.33 | -0.15 | -0.05 | 0.82 | 0.07 | -0.05 | 0.00 |
| $\mathbf{2}$ | -0.31 | -0.06 | 0.35 | 0.12 | 0.11 | -0.19 | 0.11 | 0.03 |
| $\mathbf{3}$ | 0.04 | -0.06 | -0.26 | -0.16 | 0.18 | -0.08 | -0.01 | -0.12 |
| $\mathbf{4}$ | -0.18 | 0.25 | 0.06 | 0.04 | -0.19 | -0.22 | -0.02 | -0.12 |
| $\mathbf{5}$ | 0.31 | -0.03 | 0.51 | 0.02 | -0.05 | -0.32 | 0.00 | -0.06 |
| $\mathbf{6}$ | 0.11 | 0.15 | 0.39 | 0.12 | -0.07 | -0.10 | 0.26 | -0.07 |
| $\mathbf{7}$ | 0.34 | 0.08 | 0.28 | 0.03 | -0.25 | -0.23 | 0.01 | 0.33 |
| $\mathbf{8}$ | 0.65 | 0.39 | 0.33 | 0.04 | 0.06 | -0.63 | -0.14 | 0.31 |
| $\mathbf{9}$ | 0.43 | -0.34 | 0.55 | -0.01 | -0.34 | -0.42 | -0.47 | 0.56 |

$\log ($ CNay-observed/CNay-predicted)

| Year\Age | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | -0.23 | -0.11 | -0.58 | -0.54 | -0.76 | -0.30 | 0.15 | -0.06 | 0.81 | 0.45 |
| $\mathbf{4}$ | 0.41 | -0.06 | 0.13 | 0.00 | 0.23 | 0.03 | 0.21 | 0.24 | 0.17 | 0.54 |
| $\mathbf{5}$ | -0.07 | 0.06 | -0.06 | -0.10 | 0.13 | -0.03 | -0.16 | -0.02 | -0.31 | -0.25 |
| $\mathbf{6}$ | -0.06 | 0.00 | -0.13 | 0.12 | -0.13 | 0.00 | -0.04 | -0.06 | -0.19 | -0.31 |
| $\mathbf{7}$ | -0.05 | 0.08 | 0.14 | 0.01 | -0.04 | 0.07 | 0.06 | 0.05 | -0.24 | -0.07 |
| $\mathbf{8}$ | -0.05 | 0.04 | 0.06 | 0.34 | -0.05 | 0.08 | -0.12 | -0.02 | -0.08 | 0.11 |
| $\mathbf{9}$ | 0.01 | -0.10 | -0.04 | 0.16 | -0.37 | -0.18 | -0.12 | -0.20 | 0.07 | -0.17 |
| $\mathbf{1 0}$ | 0.03 | -0.19 | -0.15 | 0.04 | -0.42 | -0.35 | -0.24 | -0.29 | 0.49 | -0.14 |
| $\mathbf{1 1}$ | -0.11 | -0.13 | -0.30 | 0.43 | -0.20 | -0.04 | -0.43 | -0.81 | 0.57 | 0.53 |
| $\mathbf{1 2}$ | 0.03 | -0.03 | -0.10 | 0.47 | 0.38 | 0.07 | 0.19 | -0.70 | 0.11 | 0.74 |
| $\mathbf{1 3}$ | 0.03 | -0.26 | 0.22 | 0.93 | 0.26 | -0.34 | -0.58 | -0.37 | -0.55 | 0.94 |
| $\mathbf{1 4}$ | 0.01 | -0.07 | 0.52 | 0.86 | -0.99 | 0.65 | 0.30 | 0.19 | 0.93 | 0.31 |


| Year\Age | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | -0.29 | -0.77 | -0.82 | -0.36 | -0.04 | 0.13 | -0.35 |
| $\mathbf{4}$ | 0.03 | 0.00 | -0.25 | -0.09 | -0.23 | 0.04 | 0.00 |
| $\mathbf{5}$ | -0.29 | -0.18 | -0.07 | -0.03 | 0.00 | -0.15 | -0.10 |
| $\mathbf{6}$ | 0.02 | -0.18 | 0.05 | 0.05 | -0.01 | -0.02 | 0.07 |
| $\mathbf{7}$ | 0.14 | 0.07 | 0.13 | 0.14 | 0.01 | 0.08 | 0.12 |
| $\mathbf{8}$ | 0.17 | 0.38 | 0.00 | 0.07 | 0.19 | -0.11 | 0.23 |
| $\mathbf{9}$ | 0.11 | 0.28 | 0.30 | -0.22 | 0.07 | 0.16 | -0.05 |
| $\mathbf{1 0}$ | 0.20 | 0.33 | 0.08 | -0.33 | -0.25 | 0.18 | 0.22 |
| $\mathbf{1 1}$ | -0.33 | 0.45 | 0.04 | -0.45 | -0.25 | -0.27 | 0.02 |
| $\mathbf{1 2}$ | -0.29 | -0.57 | 0.26 | -0.66 | -0.14 | -0.22 | -0.75 |
| $\mathbf{1 3}$ | 0.80 | -0.10 | 0.23 | -0.70 | -0.45 | -0.87 | -0.14 |
| $\mathbf{1 4}$ | 0.37 | 0.59 | -0.50 | -0.91 | -0.26 | -0.34 | -1.61 |

Table 3.3.13. (Continued)
Summary:

| Year | F5-10 | SSB | Bio4+ | N3 |
| :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 8 5}$ | 0.67 | 265 | 902 | 146 |
| $\mathbf{1 9 8 6}$ | 0.81 | 256 | 837 | 344 |
| $\mathbf{1 9 8 7}$ | 0.89 | 242 | 1031 | 268 |
| $\mathbf{1 9 8 8}$ | 0.82 | 198 | 1039 | 170 |
| $\mathbf{1 9 8 9}$ | 0.72 | 269 | 1006 | 94 |
| $\mathbf{1 9 9 0}$ | 0.70 | 342 | 844 | 135 |
| $\mathbf{1 9 9 1}$ | 0.84 | 231 | 716 | 106 |
| $\mathbf{1 9 9 2}$ | 0.92 | 250 | 571 | 173 |
| $\mathbf{1 9 9 3}$ | 0.92 | 229 | 613 | 120 |
| $\mathbf{1 9 9 4}$ | 0.65 | 258 | 573 | 74 |
| $\mathbf{1 9 9 5}$ | 0.53 | 322 | 550 | 151 |
| $\mathbf{1 9 9 6}$ | 0.50 | 266 | 660 | 172 |
| $\mathbf{1 9 9 7}$ | 0.50 | 355 | 783 | 89 |
| $\mathbf{1 9 9 8}$ | 0.58 | 348 | 720 | 166 |
| $\mathbf{1 9 9 9}$ | 0.69 | 329 | 729 | 63 |
| $\mathbf{2 0 0 0}$ | 0.72 | 250 | 571 | 184 |
| $\mathbf{2 0 0 1}$ | 0.71 | 334 | 674 | 169 |
| $\mathbf{2 0 0 2}$ | 0.58 | 354 | 717 | 174 |
| $\mathbf{2 0 0 3}$ |  | 396 | 795 | 201 |

Table 3.3.14 COLERAINE - diagnostic and results

Input data and estimated parameters:
Data: Catch-at-age 1971-2002 and spring trawl survey indices 1985-2003.

| **Dimensioning_Parameters** |  |
| :--- | ---: |
| StartYear | 1971 |
| EndYear | 2003 |
| Nsexes | 1 |
| Nages | 10 |
| Nmethods | 1 |
| NCPUEindex | 1 |
| Nsurveyindex | 1 |
| First_length | 1 |
| Length_class_increment | 1 |
| Number_of_length_classes | 1 |


| **Likelihoods** |  |
| :--- | ---: |
| CPUE | 0 |
| Survey_Index | 6.63272 |
| C@A_Commercial | -412.003 |
| C@A_survey | -254.033 |
| C@L_Commercial | 0 |
| C@L_no_sex_data_survey | 0 |
| C@L_data_survey | 0 |
| Prior_penalties | 4.94652 |


| **Parameters** |  |
| :--- | ---: |
|  |  |
| R0 | 475635 |
| h | 0.9 |


| Sex_specific |  |
| :--- | ---: |
| M | 0.2 |
| Rinit | 0.662874 |
| uinit | 0.29777 |
| plusscale | 1 |
| Method_specific |  |
| Sfullest | 5.32393 |
| log_varLest | 0.890819 |
| log_varRest | 15 |


| Method_specific_and_annual | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| errSfull | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| errvarL | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

CPUE_index_specific
log_qCPUE 0

CPUE_index_specific_and_annual qCPUEerr
q 1

| Survey_index_specific |  |
| :--- | ---: |
| log_qsurvey | -7.88422 |
| surveySfull | 4.45519 |
| log_surveyvarL | 1.61629 |
| log_surveyvarR | 15 |

## Table 3.3.14 cont'd)



Table 3.3.14 cont'd)
Estimated fishing mortality rates:

| $\mathbf{Y} / \mathbf{a}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{3}$ | 0.057 | 0.061 | 0.060 | 0.042 | 0.036 | 0.035 | 0.033 | 0.038 | 0.045 | 0.048 | 0.050 | 0.041 |
| $\mathbf{4}$ | 0.283 | 0.305 | 0.303 | 0.206 | 0.173 | 0.166 | 0.157 | 0.183 | 0.218 | 0.234 | 0.246 | 0.200 |
| $\mathbf{5}$ | 0.662 | 0.728 | 0.720 | 0.455 | 0.375 | 0.358 | 0.336 | 0.398 | 0.487 | 0.527 | 0.560 | 0.440 |
| $\mathbf{6}$ | 0.704 | 0.777 | 0.768 | 0.481 | 0.396 | 0.377 | 0.353 | 0.420 | 0.515 | 0.558 | 0.593 | 0.465 |
| $\mathbf{7}$ | 0.704 | 0.777 | 0.768 | 0.481 | 0.396 | 0.377 | 0.353 | 0.420 | 0.515 | 0.558 | 0.593 | 0.465 |
| $\mathbf{8}$ | 0.704 | 0.777 | 0.768 | 0.481 | 0.396 | 0.377 | 0.353 | 0.420 | 0.515 | 0.558 | 0.593 | 0.465 |
| $\mathbf{9}$ | 0.704 | 0.777 | 0.768 | 0.481 | 0.396 | 0.377 | 0.353 | 0.420 | 0.515 | 0.558 | 0.593 | 0.465 |
| $\mathbf{1 0 +}$ | 0.704 | 0.777 | 0.768 | 0.481 | 0.396 | 0.377 | 0.353 | 0.420 | 0.515 | 0.558 | 0.593 | 0.465 |
| $\mathbf{F 5 - 1 0}$ | $\mathbf{0 . 6 9 7}$ | $\mathbf{0 . 7 6 8}$ | $\mathbf{0 . 7 6 0}$ | $\mathbf{0 . 4 7 6}$ | $\mathbf{0 . 3 9 2}$ | $\mathbf{0 . 3 7 4}$ | $\mathbf{0 . 3 5 0}$ | $\mathbf{0 . 4 1 6}$ | $\mathbf{0 . 5 1 0}$ | $\mathbf{0 . 5 5 3}$ | $\mathbf{0 . 5 8 8}$ | $\mathbf{0 . 4 6 1}$ |

## Estimated stock in numbers:

| $\mathbf{Y} / \mathbf{a}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | 97.655 | 173.406 | 154.780 | 68.823 | 161.698 | 201.626 | 75.361 | 151.433 | 44.450 | 177.419 | 170.582 | 175.601 |
| $\mathbf{4}$ | 100.481 | 75.547 | 133.613 | 119.317 | 54.003 | 127.672 | 159.420 | 59.698 | 119.350 | 34.796 | 138.482 | 132.845 |
| $\mathbf{5}$ | 39.925 | 62.012 | 45.582 | 80.831 | 79.530 | 37.178 | 88.524 | 111.593 | 40.717 | 78.562 | 22.552 | 88.663 |
| $\mathbf{6}$ | 46.645 | 16.863 | 24.510 | 18.160 | 41.990 | 44.740 | 21.275 | 51.810 | 61.374 | 20.488 | 37.965 | 10.549 |
| $\mathbf{7}$ | 41.884 | 18.888 | 6.351 | 9.312 | 9.195 | 23.148 | 25.117 | 12.233 | 27.882 | 30.030 | 9.598 | 17.173 |
| $\mathbf{8}$ | 19.456 | 16.961 | 7.114 | 2.413 | 4.715 | 5.069 | 12.995 | 14.442 | 6.583 | 13.643 | 14.069 | 4.342 |
| $\mathbf{9}$ | 2.867 | 7.878 | 6.388 | 2.703 | 1.222 | 2.599 | 2.846 | 7.472 | 7.772 | 3.221 | 6.391 | 6.364 |
| $\mathbf{1 0 +}$ | 3.444 | 2.556 | 3.930 | 3.920 | 3.353 | 2.522 | 2.875 | 3.289 | 5.791 | 6.637 | 4.618 | 4.980 |

## Residuals

## Log(Cay-observed/Cay-predicted)

| $\mathbf{Y} / \mathbf{a}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | 0.24 | 0.18 | -0.41 | -0.10 | -0.13 | 0.05 | 0.51 | 0.23 | 0.83 | 0.85 | 0.69 |
| $\mathbf{4}$ | 0.19 | -0.04 | 0.04 | -0.18 | 0.14 | 0.05 | 0.06 | 0.14 | -0.04 | 0.16 | 0.12 |
| $\mathbf{5}$ | -0.16 | -0.17 | 0.04 | -0.06 | -0.12 | -0.29 | -0.17 | -0.14 | -0.42 | -0.33 | -0.33 |
| $\mathbf{6}$ | 0.35 | 0.16 | 0.03 | 0.68 | 0.12 | 0.24 | -0.04 | 0.28 | 0.02 | 0.09 | 0.01 |
| $\mathbf{7}$ | 0.08 | 0.40 | 0.33 | 0.27 | 0.38 | 0.05 | 0.21 | 0.04 | 0.06 | 0.29 | 0.38 |
| $\mathbf{8}$ | 0.07 | -0.07 | 0.24 | 0.45 | -0.11 | 0.07 | -0.39 | -0.05 | -0.30 | 0.39 | 0.10 |
| $\mathbf{9}$ | -0.45 | -0.26 | -0.33 | 0.21 | -0.62 | -0.71 | -0.43 | -0.70 | -0.22 | -0.46 | -0.18 |
| $\mathbf{1 0 +}$ | -0.92 | -1.16 | -0.99 | -0.45 | -0.95 | -1.14 | -1.30 | -1.32 | -0.38 | -0.47 | -0.69 |


| $\mathbf{Y} / \mathbf{a}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | -0.08 | -0.14 | -0.32 | 0.33 | 0.29 | 0.36 | -0.04 |
| $\mathbf{4}$ | -0.10 | -0.13 | -0.09 | -0.11 | -0.05 | -0.13 | -0.12 |
| $\mathbf{5}$ | -0.25 | -0.17 | -0.20 | 0.04 | -0.05 | 0.02 | -0.12 |
| $\mathbf{6}$ | 0.04 | 0.27 | 0.30 | 0.07 | 0.30 | 0.19 | 0.70 |
| $\mathbf{7}$ | 0.43 | 0.20 | 0.51 | 0.22 | -0.07 | 0.32 | 0.48 |
| $\mathbf{8}$ | 0.47 | 0.64 | -0.08 | 0.13 | 0.01 | -0.42 | 0.46 |
| $\mathbf{9}$ | 0.19 | 0.51 | 0.31 | -0.67 | -0.17 | -0.30 | -0.42 |
| $\mathbf{1 0 +}$ | -0.49 | 0.02 | -0.18 | -0.83 | -1.01 | -0.90 | -0.22 |

Table 3.3.14 (Continued)
Log(Uay-observed/Uay-predicted)

| $\mathbf{Y} / \mathbf{a}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | -0.49 | 0.41 | -0.13 | -0.79 | -0.38 | -0.27 | -0.56 | -1.31 | -0.57 | 0.64 | -0.96 |
| $\mathbf{2}$ | 0.40 | 0.22 | 0.10 | -0.51 | -0.23 | 0.08 | -0.24 | 0.07 | -0.40 | -0.07 | 0.29 |
| $\mathbf{3}$ | -0.28 | 0.06 | 0.19 | 0.34 | -0.02 | 0.03 | -0.14 | 0.05 | 0.08 | -0.30 | -0.22 |
| $\mathbf{4}$ | -0.23 | -0.47 | -0.20 | -0.04 | 0.26 | -0.23 | -0.02 | -0.07 | 0.01 | -0.14 | -0.52 |
| $\mathbf{5}$ | -0.02 | -0.30 | -0.10 | -0.06 | -0.09 | -0.07 | 0.18 | -0.06 | 0.04 | 0.01 | 0.03 |
| $\mathbf{6}$ | 0.15 | 0.14 | -0.02 | -0.33 | -0.03 | 0.19 | 0.19 | 0.33 | 0.37 | 0.17 | 0.37 |
| $\mathbf{7}$ | 0.20 | 0.00 | 0.23 | 0.05 | -0.21 | 0.13 | 0.45 | 0.08 | 0.18 | 0.37 | 0.34 |
| $\mathbf{8}$ | -0.19 | -0.51 | -0.08 | 0.36 | -0.61 | -0.20 | -0.39 | 0.08 | -0.13 | 0.08 | 0.24 |
| $\mathbf{9}$ | -0.51 | -0.88 | -0.66 | -0.76 | -0.59 | -0.58 | -0.14 | -0.53 | -0.43 | -0.41 | -0.02 |
| $\mathbf{1 0 +}$ | -1.04 | -1.24 | -0.71 | -1.27 | -1.14 | -0.76 | -0.50 | -0.74 | -0.77 | -0.53 | -0.63 |


| $\mathbf{Y} / \mathbf{a}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | -0.55 | -0.60 | -0.24 | -0.18 | 0.62 | 0.21 | -1.20 | 0.16 |
| $\mathbf{2}$ | -0.44 | 0.12 | -0.20 | 0.30 | 0.07 | 0.08 | 0.18 | -0.47 |
| $\mathbf{3}$ | 0.05 | -0.26 | -0.32 | -0.43 | 0.14 | 0.06 | -0.04 | -0.06 |
| $\mathbf{4}$ | -0.24 | 0.03 | -0.39 | 0.00 | -0.68 | -0.09 | -0.17 | -0.12 |
| $\mathbf{5}$ | 0.12 | -0.08 | 0.29 | -0.15 | -0.07 | -0.35 | 0.09 | 0.07 |
| $\mathbf{6}$ | 0.07 | 0.23 | 0.28 | 0.04 | -0.01 | 0.26 | 0.59 | 0.22 |
| $\mathbf{7}$ | 0.68 | -0.02 | 0.25 | 0.07 | -0.41 | 0.12 | 0.26 | 0.84 |
| $\mathbf{8}$ | 0.90 | 0.36 | -0.14 | -0.04 | -0.30 | -0.79 | -0.08 | 0.49 |
| $\mathbf{9}$ | 0.28 | -0.59 | -0.02 | -0.78 | -0.95 | -0.92 | -1.19 | 0.39 |
| $\mathbf{1 0 +}$ | -0.32 | -0.42 | -0.52 | -0.90 | -1.82 | -1.13 | -1.79 | -1.57 |

Total survey index residuals

| Year | Obs | Fit | $\boldsymbol{\operatorname { l o g } ( \text { Obs/Fit } )}$ |
| :---: | :---: | :---: | :---: |
| $\mathbf{1 9 8 5}$ | 548.4 | 384.1 | 0.4 |
| $\mathbf{1 9 8 6}$ | 402.2 | 419.8 | 0.0 |
| $\mathbf{1 9 8 7}$ | 493.1 | 464.8 | 0.1 |
| $\mathbf{1 9 8 8}$ | 561.3 | 459.5 | 0.2 |
| $\mathbf{1 9 8 9}$ | 522.0 | 424.9 | 0.2 |
| $\mathbf{1 9 9 0}$ | 307.0 | 346.7 | -0.1 |
| $\mathbf{1 9 9 1}$ | 291.9 | 285.9 | 0.0 |
| $\mathbf{1 9 9 2}$ | 203.8 | 252.2 | -0.2 |
| $\mathbf{1 9 9 3}$ | 220.3 | 259.7 | -0.2 |
| $\mathbf{1 9 9 4}$ | 177.8 | 237.6 | -0.3 |
| $\mathbf{1 9 9 5}$ | 224.2 | 256.4 | -0.1 |
| $\mathbf{1 9 9 6}$ | 325.9 | 313.9 | 0.0 |
| $\mathbf{1 9 9 7}$ | 347.0 | 329.7 | 0.1 |
| $\mathbf{1 9 9 8}$ | 422.6 | 322.6 | 0.3 |
| $\mathbf{1 9 9 9}$ | 282.2 | 288.1 | 0.0 |
| $\mathbf{2 0 0 0}$ | 251.9 | 263.4 | 0.0 |
| $\mathbf{2 0 0 1}$ | 236.1 | 292.2 | -0.2 |
| $\mathbf{2 0 0 2}$ | 310.0 | 298.5 | 0.0 |
| $\mathbf{2 0 0 3}$ | 344.5 | 342.3 | 0.0 |

Table 3.3.14 (Continued)

## Summary

| Year | F5-10 | Bio4+ | N3 |
| :---: | :---: | :---: | :---: |
| $\mathbf{1 9 8 5}$ | 0.49 | 1020 | 124 |
| $\mathbf{1 9 8 6}$ | 0.65 | 900 | 354 |
| $\mathbf{1 9 8 7}$ | 0.70 | 1090 | 340 |
| $\mathbf{1 9 8 8}$ | 0.57 | 1190 | 185 |
| $\mathbf{1 9 8 9}$ | 0.46 | 1200 | 78 |
| $\mathbf{1 9 9 0}$ | 0.55 | 913 | 129 |
| $\mathbf{1 9 9 1}$ | 0.70 | 763 | 98 |
| $\mathbf{1 9 9 2}$ | 0.77 | 599 | 173 |
| $\mathbf{1 9 9 3}$ | 0.76 | 631 | 155 |
| $\mathbf{1 9 9 4}$ | 0.48 | 642 | 69 |
| $\mathbf{1 9 9 5}$ | 0.39 | 613 | 162 |
| $\mathbf{1 9 9 6}$ | 0.37 | 738 | 202 |
| $\mathbf{1 9 9 7}$ | 0.35 | 906 | 75 |
| $\mathbf{1 9 9 8}$ | 0.42 | 829 | 151 |
| $\mathbf{1 9 9 9}$ | 0.51 | 817 | 44 |
| $\mathbf{2 0 0 0}$ | 0.55 | 623 | 177 |
| $\mathbf{2 0 0 1}$ | 0.59 | 701 | 171 |
| $\mathbf{2 0 0 2}$ | 0.46 | 737 | 176 |
| $\mathbf{2 0 0 3}$ |  | 809 | 229 |

Table 3.3.15 Comparison of the results from the variuos methods.

Estimated fishing mortality rate in 2002:

| Age | XSA | TSA | AD-CAM | EX-CAM | ADAPT | Coleraine |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{3}$ | 0.04 |  | 0.04 | 0.06 | 0.04 | 0.04 |
| $\mathbf{4}$ | 0.18 | 0.17 | 0.18 | 0.17 | 0.18 | 0.18 |
| $\mathbf{5}$ | 0.35 | 0.33 | 0.36 | 0.37 | 0.34 | 0.38 |
| $\mathbf{6}$ | 0.55 | 0.50 | 0.61 | 0.55 | 0.56 | 0.40 |
| $\mathbf{7}$ | 0.69 | 0.68 | 0.75 | 0.62 | 0.69 | 0.40 |
| $\mathbf{8}$ | 0.80 | 0.80 | 0.88 | 0.65 | 0.87 | 0.40 |
| $\mathbf{9}$ | 0.92 | 0.82 | 0.95 | 0.65 | 1.25 | 0.40 |
| $\mathbf{1 0}$ | 0.99 | 0.83 | 1.01 | 0.65 | 2.19 | 0.40 |
| $\mathbf{1 1}$ | 1.00 | 0.82 | 0.94 | 0.65 | 2.31 |  |
| $\mathbf{1 2}$ | 1.05 |  | 0.96 | 0.65 | 0.37 |  |
| $\mathbf{1 3}$ | 1.01 |  | 1.57 | 0.65 | 3.54 |  |
| $\mathbf{1 4}$ | 1.02 |  | 1.57 | 0.65 | 3.54 |  |
| $\mathbf{F}(\mathbf{5 - 1 0})$ | $\mathbf{0 . 7 2}$ | $\mathbf{0 . 6 6}$ | $\mathbf{0 . 7 6}$ | $\mathbf{0 . 5 8}$ | $\mathbf{0 . 9 8}$ | $\mathbf{0 . 4 0}$ |

Estimated stock in numbers in 2003:

| Age | XSA | TSA | AD-CAM | EX-CAM | ADAPT | Coleraine |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | 198.028 | 187.000 | 207.040 | 200.922 | 206.202 | 229.330 |
| $\mathbf{4}$ | 136.855 | 126.372 | 130.668 | 134.571 | 132.559 | 137.941 |
| $\mathbf{5}$ | 86.829 | 90.966 | 86.862 | 89.380 | 86.201 | 89.064 |
| $\mathbf{6}$ | 52.048 | 58.054 | 53.290 | 52.837 | 54.105 | 46.737 |
| $\mathbf{7}$ | 8.548 | 8.785 | 7.449 | 7.890 | 8.351 | 5.425 |
| $\mathbf{8}$ | 8.169 | 8.225 | 7.390 | 8.194 | 8.111 | 8.832 |
| $\mathbf{9}$ | 1.639 | 1.488 | 1.349 | 1.736 | 1.451 | 2.233 |
| $\mathbf{1 0}$ | 0.812 | 0.956 | 0.833 | 1.385 | 0.490 | 5.834 |
| $\mathbf{1 1}$ | 0.369 | 0.441 | 0.325 | 0.534 | 0.079 |  |
| $\mathbf{1 2}$ | 0.065 |  | 0.065 | 0.117 | 0.012 |  |
| $\mathbf{1 3}$ | 0.019 |  | 0.035 | 0.082 | 0.081 |  |
| $\mathbf{1 4}$ | 0.021 |  | 0.010 | 0.046 | 0.001 |  |

Recruitment:

| Year class | RCT3 | XSA | AD-CAM | EX-CAM | Coleraine | ADAPT | TSA |
| :--- | :--- | :--- | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 9 8}$ | 167 | 167 | 166 | 169 | 171 | 166 | 171 |
| $\mathbf{1 9 9 9}$ | 177 | 173 | 167 | 174 | 176 | 168 | 164 |
| $\mathbf{2 0 0 0}$ | 202 | 198 | 207 | 201 | 229 | 206 | 187 |
| $\mathbf{2 0 0 1}$ | 69 |  | 69 | 66 | 59 | 72 | 75 |
| $\mathbf{2 0 0 2}$ | 195 |  | 195 | 201 |  | 192 | 187 |

Estimated stock in weight (4+, Thous. tonnes) in 1990-2003

| Year | XSA | TSA | AD-CAM | EX-CAM | ADAPT | Coleraine |
| :---: | :---: | :---: | :---: | :---: | :---: | ---: |
| $\mathbf{1 9 9 0}$ | 847 | 821 | 837 | 844 | 845 | 913 |
| $\mathbf{1 9 9 1}$ | 711 | 687 | 676 | 716 | 709 | 763 |
| $\mathbf{1 9 9 2}$ | 554 | 532 | 539 | 571 | 553 | 599 |
| $\mathbf{1 9 9 3}$ | 588 | 575 | 572 | 613 | 587 | 631 |
| $\mathbf{1 9 9 4}$ | 584 | 577 | 581 | 573 | 584 | 642 |
| $\mathbf{1 9 9 5}$ | 562 | 558 | 548 | 550 | 561 | 613 |
| $\mathbf{1 9 9 6}$ | 688 | 676 | 656 | 660 | 681 | 738 |
| $\mathbf{1 9 9 7}$ | 805 | 793 | 794 | 783 | 800 | 906 |
| $\mathbf{1 9 9 8}$ | 735 | 722 | 720 | 720 | 729 | 829 |
| $\mathbf{1 9 9 9}$ | 738 | 726 | 730 | 729 | 747 | 817 |
| $\mathbf{2 0 0 0}$ | 575 | 564 | 559 | 571 | 571 | 623 |
| $\mathbf{2 0 0 1}$ | 669 | 684 | 663 | 674 | 669 | 701 |
| $\mathbf{2 0 0 2}$ | 706 | 738 | 704 | 717 | 708 | 737 |
| $\mathbf{2 0 0 3}$ | $\mathbf{7 8 3}$ | $\mathbf{7 9 9}$ | $\mathbf{7 6 6}$ | $\mathbf{7 9 5}$ | $\mathbf{7 7 2}$ | $\mathbf{8 0 9}$ |

Table 3.3.16 Cod at Iceland. Division Va. Resulting fishing mortality using final F from
AD-CAM using catch-at-age and spring trawl survey indices

| Year/Age | $\mathbf{1 9 7 4}$ | $\mathbf{1 9 7 5}$ | $\mathbf{1 9 7 6}$ | $\mathbf{1 9 7 7}$ | $\mathbf{1 9 7 8}$ | $\mathbf{1 9 7 9}$ | $\mathbf{1 9 8 0}$ | $\mathbf{1 9 8 1}$ | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | 0.106 | 0.114 | 0.065 | 0.027 | 0.029 | 0.032 | 0.030 | 0.020 | 0.025 | 0.022 |
| $\mathbf{4}$ | 0.384 | 0.303 | 0.263 | 0.179 | 0.171 | 0.180 | 0.173 | 0.151 | 0.196 | 0.144 |
| $\mathbf{5}$ | 0.505 | 0.527 | 0.386 | 0.337 | 0.261 | 0.246 | 0.328 | 0.384 | 0.408 | 0.408 |
| $\mathbf{6}$ | 0.524 | 0.556 | 0.555 | 0.412 | 0.359 | 0.354 | 0.398 | 0.462 | 0.543 | 0.578 |
| $\mathbf{7}$ | 0.739 | 0.685 | 0.671 | 0.707 | 0.548 | 0.508 | 0.522 | 0.602 | 0.623 | 0.701 |
| $\mathbf{8}$ | 0.875 | 0.880 | 0.868 | 0.748 | 0.678 | 0.579 | 0.594 | 0.676 | 0.875 | 0.843 |
| $\mathbf{9}$ | 0.800 | 0.931 | 0.902 | 0.867 | 0.687 | 0.617 | 0.631 | 0.783 | 0.976 | 0.911 |
| $\mathbf{1 0}$ | 0.985 | 0.981 | 0.937 | 0.815 | 0.697 | 0.641 | 0.664 | 0.779 | 0.830 | 0.854 |
| $\mathbf{1 1}$ | 0.974 | 0.955 | 0.881 | 0.796 | 0.750 | 0.716 | 0.702 | 0.705 | 0.665 | 0.653 |
| $\mathbf{1 2}$ | 0.757 | 0.756 | 0.733 | 0.707 | 0.694 | 0.687 | 0.686 | 0.683 | 0.656 | 0.647 |
| $\mathbf{1 3}$ | 0.397 | 0.399 | 0.402 | 0.405 | 0.414 | 0.424 | 0.433 | 0.443 | 0.451 | 0.457 |
| $\mathbf{1 4}$ | 0.397 | 0.399 | 0.402 | 0.405 | 0.414 | 0.424 | 0.433 | 0.443 | 0.451 | 0.457 |
| $\mathbf{F} \mathbf{5 - 1 0}$ | $\mathbf{0 . 7 3 8}$ | $\mathbf{0 . 7 6 0}$ | $\mathbf{0 . 7 2 0}$ | $\mathbf{0 . 6 4 7}$ | $\mathbf{0 . 5 3 8}$ | $\mathbf{0 . 4 9 1}$ | $\mathbf{0 . 5 2 3}$ | $\mathbf{0 . 6 1 4}$ | $\mathbf{0 . 7 0 9}$ | $\mathbf{0 . 7 1 6}$ |


| Year/Age | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | 0.046 | 0.053 | 0.063 | 0.044 | 0.043 | 0.038 | 0.052 | 0.086 | 0.088 | 0.149 |
| $\mathbf{4}$ | 0.196 | 0.269 | 0.249 | 0.289 | 0.211 | 0.272 | 0.251 | 0.318 | 0.387 | 0.312 |
| $\mathbf{5}$ | 0.351 | 0.397 | 0.559 | 0.545 | 0.496 | 0.383 | 0.487 | 0.548 | 0.590 | 0.505 |
| $\mathbf{6}$ | 0.557 | 0.605 | 0.691 | 0.768 | 0.752 | 0.645 | 0.620 | 0.797 | 0.843 | 0.734 |
| $\mathbf{7}$ | 0.663 | 0.724 | 0.853 | 0.930 | 0.921 | 0.804 | 0.834 | 0.906 | 1.022 | 0.865 |
| $\mathbf{8}$ | 0.856 | 0.820 | 0.932 | 1.047 | 1.166 | 0.960 | 0.870 | 0.884 | 1.002 | 1.072 |
| $\mathbf{9}$ | 0.787 | 0.802 | 0.834 | 0.924 | 0.975 | 0.851 | 0.812 | 0.815 | 0.817 | 1.019 |
| $\mathbf{1 0}$ | 0.770 | 0.748 | 0.744 | 0.744 | 0.762 | 0.735 | 0.752 | 0.781 | 0.789 | 0.853 |
| $\mathbf{1 1}$ | 0.643 | 0.644 | 0.651 | 0.657 | 0.681 | 0.680 | 0.688 | 0.696 | 0.702 | 0.720 |
| $\mathbf{1 2}$ | 0.633 | 0.628 | 0.631 | 0.635 | 0.647 | 0.652 | 0.653 | 0.656 | 0.653 | 0.655 |
| $\mathbf{1 3}$ | 0.462 | 0.465 | 0.465 | 0.468 | 0.469 | 0.466 | 0.464 | 0.463 | 0.463 | 0.463 |
| $\mathbf{1 4}$ | 0.462 | 0.465 | 0.465 | 0.468 | 0.469 | 0.466 | 0.464 | 0.463 | 0.463 | 0.463 |
| $\mathbf{F} \mathbf{5 - 1 0}$ | $\mathbf{0 . 6 6 4}$ | $\mathbf{0 . 6 8 3}$ | $\mathbf{0 . 7 6 9}$ | $\mathbf{0 . 8 2 6}$ | $\mathbf{0 . 8 4 6}$ | $\mathbf{0 . 7 3 0}$ | $\mathbf{0 . 7 2 9}$ | $\mathbf{0 . 7 8 8}$ | $\mathbf{0 . 8 4 4}$ | $\mathbf{0 . 8 4 1}$ |


| Year/Age | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | 0.101 | 0.067 | 0.037 | 0.026 | 0.029 | 0.051 | 0.066 | 0.069 | 0.044 |
| $\mathbf{4}$ | 0.294 | 0.185 | 0.147 | 0.146 | 0.146 | 0.186 | 0.193 | 0.217 | 0.177 |
| $\mathbf{5}$ | 0.323 | 0.359 | 0.274 | 0.265 | 0.367 | 0.406 | 0.439 | 0.399 | 0.356 |
| $\mathbf{6}$ | 0.561 | 0.394 | 0.454 | 0.445 | 0.526 | 0.696 | 0.657 | 0.644 | 0.608 |
| $\mathbf{7}$ | 0.732 | 0.525 | 0.531 | 0.614 | 0.664 | 0.723 | 0.869 | 0.791 | 0.747 |
| $\mathbf{8}$ | 0.927 | 0.566 | 0.614 | 0.680 | 0.765 | 0.774 | 0.860 | 0.914 | 0.875 |
| $\mathbf{9}$ | 0.953 | 0.575 | 0.626 | 0.725 | 0.834 | 0.844 | 0.868 | 0.926 | 0.948 |
| $\mathbf{1 0}$ | 0.854 | 0.785 | 0.800 | 0.838 | 0.875 | 0.898 | 0.943 | 0.986 | 1.007 |
| $\mathbf{1 1}$ | 0.720 | 0.857 | 0.871 | 0.891 | 0.904 | 0.905 | 0.919 | 0.936 | 0.939 |
| $\mathbf{1 2}$ | 0.657 | 0.956 | 0.952 | 0.956 | 0.964 | 0.966 | 0.971 | 0.966 | 0.957 |
| $\mathbf{1 3}$ | 0.463 | 1.580 | 1.580 | 1.580 | 1.580 | 1.578 | 1.576 | 1.574 | 1.574 |
| $\mathbf{1 4}$ | 0.463 | 1.580 | 1.580 | 1.580 | 1.580 | 1.578 | 1.576 | 1.574 | 1.574 |
| $\mathbf{F}(\mathbf{5 - 1 0}$ | $\mathbf{0 . 7 2 5}$ | $\mathbf{0 . 5 3 4}$ | $\mathbf{0 . 5 5 0}$ | $\mathbf{0 . 5 9 4}$ | $\mathbf{0 . 6 7 2}$ | $\mathbf{0 . 7 2 3}$ | $\mathbf{0 . 7 7 3}$ | $\mathbf{0 . 7 7 7}$ | $\mathbf{0 . 7 5 7}$ |

Table 3.3.17 Cod at Iceland. Division Va. Resulting Stock in numbers using final F from AD-CAM using catch-at-age and spring trawl survey indices.

| $\mathbf{Y} / \mathbf{A}$ | $\mathbf{1 9 7 4}$ | $\mathbf{1 9 7 5}$ | $\mathbf{1 9 7 6}$ | $\mathbf{1 9 7 7}$ | $\mathbf{1 9 7 8}$ | $\mathbf{1 9 7 9}$ | $\mathbf{1 9 8 0}$ | $\mathbf{1 9 8 1}$ | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | 584.188 | 208.152 | 334.404 | 363.502 | 218.983 | 215.482 | 199.628 | 335.005 | 210.231 | 203.229 |
| $\mathbf{2}$ | 318.613 | 478.293 | 170.420 | 273.787 | 297.611 | 179.288 | 176.422 | 163.442 | 274.279 | 172.122 |
| $\mathbf{3}$ | 177.179 | 260.858 | 391.593 | 139.528 | 224.158 | 243.663 | 146.789 | 144.442 | 133.815 | 224.560 |
| $\mathbf{4}$ | 202.658 | 130.455 | 190.475 | 300.475 | 111.196 | 178.197 | 193.282 | 116.615 | 115.960 | 106.842 |
| $\mathbf{5}$ | 61.236 | 112.971 | 78.895 | 119.871 | 205.660 | 76.694 | 121.807 | 133.118 | 82.057 | 78.064 |
| $\mathbf{6}$ | 43.313 | 30.262 | 54.619 | 43.907 | 70.087 | 129.719 | 49.100 | 71.843 | 74.199 | 44.661 |
| $\mathbf{7}$ | 18.684 | 21.002 | 14.209 | 25.666 | 23.815 | 40.079 | 74.560 | 26.996 | 37.059 | 35.307 |
| $\mathbf{8}$ | 10.658 | 7.307 | 8.668 | 5.945 | 10.363 | 11.278 | 19.737 | 60.535 | 12.104 | 16.281 |
| $\mathbf{9}$ | 3.555 | 3.637 | 2.483 | 2.979 | 2.304 | 4.307 | 5.174 | 8.921 | 15.085 | 4.133 |
| $\mathbf{1 0}$ | 4.366 | 1.308 | 1.174 | 0.824 | 1.025 | 0.949 | 1.903 | 2.253 | 3.338 | 4.652 |
| $\mathbf{1 1}$ | 2.181 | 1.335 | 0.401 | 0.377 | 0.299 | 0.418 | 0.409 | 0.802 | 0.847 | 1.191 |
| $\mathbf{1 2}$ | 0.596 | 0.674 | 0.420 | 0.136 | 0.139 | 0.116 | 0.167 | 0.166 | 0.324 | 0.356 |
| $\mathbf{1 3}$ | 0.230 | 0.229 | 0.259 | 0.165 | 0.055 | 0.057 | 0.048 | 0.069 | 0.069 | 0.138 |
| $\mathbf{1 4}$ | 0.037 | 0.126 | 0.126 | 0.142 | 0.090 | 0.030 | 0.030 | 0.025 | 0.036 | 0.036 |


| $\mathbf{Y} / \mathbf{A}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | 511.615 | 449.381 | 270.692 | 128.808 | 192.275 | 160.707 | 247.115 | 215.691 | 110.888 | 221.097 |
| $\mathbf{2}$ | 166.389 | 418.875 | 367.922 | 221.624 | 105.459 | 157.422 | 131.576 | 202.320 | 176.593 | 90.788 |
| $\mathbf{3}$ | 140.922 | 136.228 | 342.946 | 301.229 | 181.450 | 86.342 | 128.886 | 107.725 | 165.646 | 144.582 |
| $\mathbf{4}$ | 179.829 | 110.175 | 105.733 | 263.611 | 236.111 | 142.364 | 68.037 | 100.156 | 80.937 | 124.142 |
| $\mathbf{5}$ | 75.758 | 121.086 | 68.904 | 67.500 | 161.710 | 156.565 | 88.836 | 43.326 | 59.644 | 44.984 |
| $\mathbf{6}$ | 42.498 | 43.649 | 66.679 | 32.253 | 32.036 | 80.591 | 96.856 | 44.697 | 20.515 | 27.066 |
| $\mathbf{7}$ | 20.514 | 19.929 | 19.521 | 27.355 | 12.250 | 12.368 | 34.605 | 38.486 | 16.494 | 7.232 |
| $\mathbf{8}$ | 14.346 | 8.657 | 7.907 | 6.812 | 8.835 | 3.993 | 4.531 | 12.310 | 12.739 | 4.862 |
| $\mathbf{9}$ | 5.734 | 4.989 | 3.120 | 2.548 | 1.958 | 2.253 | 1.252 | 1.554 | 4.165 | 3.831 |
| $\mathbf{1 0}$ | 1.361 | 2.137 | 1.831 | 1.110 | 0.828 | 0.604 | 0.788 | 0.455 | 0.563 | 1.506 |
| $\mathbf{1 1}$ | 1.621 | 0.516 | 0.828 | 0.712 | 0.432 | 0.316 | 0.237 | 0.304 | 0.171 | 0.210 |
| $\mathbf{1 2}$ | 0.507 | 0.698 | 0.222 | 0.353 | 0.302 | 0.179 | 0.131 | 0.098 | 0.124 | 0.069 |
| $\mathbf{1 3}$ | 0.153 | 0.221 | 0.305 | 0.097 | 0.153 | 0.130 | 0.076 | 0.056 | 0.041 | 0.053 |
| $\mathbf{1 4}$ | 0.071 | 0.079 | 0.113 | 0.157 | 0.050 | 0.079 | 0.067 | 0.039 | 0.029 | 0.021 |


| $\mathbf{Y} / \mathbf{A}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | 270.233 | 121.939 | 232.554 | 86.713 | 275.369 | 247.807 | 249.161 | 309.069 | 103.448 | 291.515 |
| $\mathbf{2}$ | 181.019 | 221.248 | 99.835 | 190.399 | 70.994 | 225.453 | 202.887 | 203.996 | 253.045 | 84.696 |
| $\mathbf{3}$ | 74.331 | 148.206 | 181.142 | 81.738 | 155.886 | 58.125 | 184.586 | 166.110 | 167.018 | 207.175 |
| $\mathbf{4}$ | 101.955 | 55.009 | 113.494 | 142.950 | 65.186 | 124.042 | 45.238 | 141.523 | 126.912 | 130.817 |
| $\mathbf{5}$ | 74.424 | 62.217 | 37.425 | 80.198 | 101.143 | 46.141 | 84.308 | 30.540 | 93.262 | 87.067 |
| $\mathbf{6}$ | 22.237 | 44.107 | 35.561 | 23.306 | 50.380 | 57.385 | 25.181 | 44.508 | 16.778 | 53.486 |
| $\mathbf{7}$ | 10.637 | 10.386 | 24.351 | 18.487 | 12.230 | 24.385 | 23.435 | 10.689 | 19.137 | 7.477 |
| $\mathbf{8}$ | 2.493 | 4.190 | 5.028 | 11.718 | 8.192 | 5.156 | 9.685 | 8.048 | 3.968 | 7.421 |
| $\mathbf{9}$ | 1.363 | 0.807 | 1.948 | 2.228 | 4.861 | 3.120 | 1.947 | 3.355 | 2.643 | 1.354 |
| $\mathbf{1 0}$ | 1.133 | 0.430 | 0.372 | 0.853 | 0.884 | 1.728 | 1.099 | 0.669 | 1.089 | 0.839 |
| $\mathbf{1 1}$ | 0.525 | 0.395 | 0.161 | 0.137 | 0.302 | 0.302 | 0.577 | 0.350 | 0.204 | 0.326 |
| $\mathbf{1 2}$ | 0.083 | 0.209 | 0.137 | 0.055 | 0.046 | 0.100 | 0.100 | 0.188 | 0.112 | 0.065 |
| $\mathbf{1 3}$ | 0.029 | 0.035 | 0.066 | 0.043 | 0.017 | 0.014 | 0.031 | 0.031 | 0.059 | 0.035 |
| $\mathbf{1 4}$ | 0.027 | 0.015 | 0.006 | 0.011 | 0.007 | 0.003 | 0.002 | 0.005 | 0.005 | 0.010 |

Table 3.3.18 Cod at Iceland. Division Va. Resulting SSB using final F from AD-CAM using catch-at-age and spring trawl survey indices.

| $\mathbf{Y} / \mathbf{A}$ | $\mathbf{1 9 7 4}$ | $\mathbf{1 9 7 5}$ | $\mathbf{1 9 7 6}$ | $\mathbf{1 9 7 7}$ | $\mathbf{1 9 7 8}$ | $\mathbf{1 9 7 9}$ | $\mathbf{1 9 8 0}$ | $\mathbf{1 9 8 1}$ | $\mathbf{1 9 8 2}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | 2.446 | 1.682 | 22.091 | 0.013 | 10.745 | 0.026 | 10.396 | 0.013 | 2.910 |
| $\mathbf{4}$ | 29.951 | 24.413 | 16.076 | 22.412 | 8.569 | 5.135 | 6.886 | 4.736 | 8.830 |
| $\mathbf{5}$ | 39.463 | 106.171 | 48.214 | 60.971 | 97.126 | 32.846 | 47.727 | 20.557 | 19.063 |
| $\mathbf{6}$ | 94.703 | 54.512 | 97.504 | 92.803 | 104.107 | 237.620 | 76.892 | 55.553 | 40.834 |
| $\mathbf{7}$ | 62.347 | 77.665 | 35.569 | 97.804 | 96.320 | 141.013 | 255.363 | 60.773 | 65.597 |
| $\mathbf{8}$ | 46.321 | 32.159 | 36.889 | 26.187 | 48.978 | 53.782 | 94.158 | 223.165 | 37.159 |
| $\mathbf{9}$ | 17.335 | 17.296 | 13.277 | 14.134 | 14.273 | 24.685 | 28.949 | 42.599 | 58.559 |
| $\mathbf{1 0}$ | 20.217 | 6.820 | 7.243 | 5.141 | 7.577 | 5.805 | 13.834 | 12.917 | 19.131 |
| $\mathbf{1 1}$ | 10.823 | 7.040 | 2.968 | 3.082 | 2.568 | 3.653 | 3.518 | 5.833 | 6.773 |
| $\mathbf{1 2}$ | 3.906 | 4.377 | 3.248 | 1.284 | 1.222 | 1.032 | 1.939 | 1.415 | 3.034 |
| $\mathbf{1 3}$ | 2.029 | 1.814 | 2.832 | 2.209 | 0.821 | 0.588 | 0.587 | 0.623 | 0.903 |
| $\mathbf{1 4}$ | 0.423 | 0.982 | 1.517 | 2.744 | 1.494 | 0.438 | 0.595 | 0.279 | 0.461 |
| Total | 329.965 | 334.932 | 287.428 | 328.784 | 393.798 | 506.622 | 540.845 | 428.463 | 263.254 |


| $\mathbf{Y} / \mathbf{A}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | 0.013 | 3.939 | 1.918 | 7.360 | 8.169 | 0.008 | 0.010 | 0.011 | 9.864 |
| $\mathbf{4}$ | 10.503 | 9.248 | 9.151 | 19.832 | 6.937 | 9.545 | 6.170 | 10.066 | 22.441 |
| $\mathbf{5}$ | 28.175 | 48.184 | 37.325 | 36.509 | 70.014 | 74.169 | 48.380 | 17.193 | 59.881 |
| $\mathbf{6}$ | 49.566 | 67.041 | 99.986 | 60.080 | 41.494 | 130.373 | 154.679 | 55.509 | 34.194 |
| $\mathbf{7}$ | 44.520 | 55.740 | 49.279 | 75.306 | 29.147 | 38.004 | 91.976 | 78.713 | 43.115 |
| $\mathbf{8}$ | 43.212 | 32.186 | 28.791 | 25.099 | 26.167 | 14.804 | 17.811 | 41.625 | 41.997 |
| $\mathbf{9}$ | 24.629 | 23.680 | 15.345 | 11.586 | 7.934 | 9.785 | 6.868 | 7.251 | 17.760 |
| $\mathbf{1 0}$ | 8.438 | 13.124 | 11.066 | 7.568 | 4.480 | 3.770 | 5.346 | 2.654 | 3.024 |
| $\mathbf{1 1}$ | 12.131 | 4.090 | 6.087 | 5.427 | 2.945 | 2.601 | 1.965 | 2.328 | 1.358 |
| $\mathbf{1 2}$ | 4.917 | 6.529 | 2.132 | 2.637 | 1.994 | 1.073 | 1.412 | 0.974 | 1.164 |
| $\mathbf{1 3}$ | 1.792 | 2.577 | 3.207 | 1.168 | 1.652 | 1.034 | 0.892 | 0.604 | 0.489 |
| $\mathbf{1 4}$ | 0.893 | 1.136 | 1.598 | 1.969 | 0.412 | 0.854 | 0.876 | 0.605 | 0.247 |
| Total | 228.791 | 267.474 | 265.884 | 254.541 | 201.344 | 286.020 | 336.386 | 217.536 | 235.534 |


| $\mathbf{Y} / \mathbf{A}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | 8.029 | 6.426 | 16.935 | 6.915 | 5.778 | 2.822 | 7.159 | 22.892 | 19.514 | 20.210 |
| $\mathbf{4}$ | 45.182 | 36.411 | 16.592 | 65.162 | 19.747 | 49.444 | 11.020 | 86.672 | 81.834 | 65.118 |
| $\mathbf{5}$ | 88.717 | 107.937 | 44.281 | 93.059 | 79.997 | 49.056 | 68.327 | 39.051 | 118.191 | 101.249 |
| $\mathbf{6}$ | 52.534 | 110.852 | 84.016 | 54.749 | 81.785 | 101.192 | 40.837 | 89.618 | 40.882 | 117.718 |
| $\mathbf{7}$ | 35.436 | 32.686 | 75.054 | 66.513 | 36.461 | 68.850 | 59.644 | 33.601 | 57.957 | 24.242 |
| $\mathbf{8}$ | 9.024 | 18.187 | 19.299 | 49.173 | 40.058 | 21.370 | 34.672 | 30.188 | 17.225 | 31.962 |
| $\mathbf{9}$ | 6.316 | 4.425 | 8.727 | 11.113 | 22.697 | 16.447 | 9.458 | 15.750 | 13.222 | 7.351 |
| $\mathbf{1 0}$ | 5.383 | 2.783 | 2.278 | 4.363 | 4.909 | 10.673 | 6.178 | 3.611 | 5.850 | 5.177 |
| $\mathbf{1 1}$ | 3.785 | 2.702 | 1.140 | 0.892 | 2.039 | 2.088 | 3.621 | 1.937 | 1.341 | 2.219 |
| $\mathbf{1 2}$ | 0.731 | 1.427 | 1.061 | 0.357 | 0.421 | 0.755 | 0.702 | 1.119 | 0.947 | 0.519 |
| $\mathbf{1 3}$ | 0.323 | 0.208 | 0.396 | 0.212 | 0.142 | 0.096 | 0.184 | 0.144 | 0.381 | 0.239 |
| $\mathbf{1 4}$ | 0.363 | 0.103 | 0.043 | 0.076 | 0.053 | 0.020 | 0.019 | 0.033 | 0.040 | 0.085 |
| Total | 255.823 | 324.147 | 269.822 | 352.584 | 294.086 | 322.814 | 241.821 | 324.613 | 357.382 | 376.092 |

Table 3.3.19 Cod at Iceland. Division Va. Resulting stock weight using final F from

AD-CAM using catch-at-age and spring trawl survey indices

| Y/A | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 186.038 | 286.944 | 528.651 | 175.666 | 288.940 | 343.078 | 204.330 | 170.442 | 134.618 | 245.893 |
| 4 | 346.545 | 230.905 | 339.046 | 574.208 | 203.822 | 348.553 | 359.891 | 192.531 | 179.738 | 170.840 |
| 5 | 148.804 | 314.059 | 209.072 | 342.352 | 602.378 | 202.624 | 332.899 | 300.847 | 184.301 | 177.597 |
| 6 | 165.457 | 113.787 | 223.940 | 178.659 | 277.195 | 518.746 | 185.010 | 236.579 | 230.313 | 134.921 |
| 7 | 97.902 | 114.463 | 72.039 | 148.271 | 136.365 | 222.357 | 392.112 | 121.023 | 157.796 | 144.617 |
| 8 | 70.984 | 48.883 | 58.334 | 39.449 | 70.528 | 76.168 | 137.783 | 352.371 | 65.193 | 89.233 |
| 9 | 25.420 | 27.534 | 20.481 | 22.894 | 20.828 | 35.745 | 41.582 | 69.038 | 100.801 | 29.130 |
| 10 | 33.882 | 11.223 | 11.278 | 8.021 | 11.138 | 8.838 | 20.422 | 21.227 | 30.511 | 37.811 |
| 11 | 17.864 | 11.760 | 4.633 | 4.407 | 3.906 | 5.491 | 5.034 | 9.122 | 10.128 | 13.115 |
| 12 | 5.829 | 6.595 | 4.805 | 1.961 | 1.666 | 1.550 | 2.891 | 2.124 | 4.615 | 4.981 |
| 13 | 2.843 | 2.309 | 3.644 | 2.888 | 1.049 | 0.770 | 0.709 | 0.863 | 1.187 | 2.188 |
| 14 | 0.550 | 1.391 | 2.034 | 3.424 | 1.923 | 0.598 | 0.581 | 0.482 | 0.602 | 0.663 |
| 4+ | 916.08 | 882.91 | 949.31 | 1326.53 | 1330.80 | 1421.44 | 1478.91 | 1306.21 | 965.19 | 805.10 |


| $\mathbf{Y} / \mathbf{A}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | 181.508 | 191.673 | 500.358 | 396.417 | 260.925 | 102.402 | 166.263 | 141.012 | 213.518 |
| $\mathbf{4}$ | 310.205 | 217.155 | 207.342 | 515.623 | 426.180 | 258.106 | 115.934 | 190.196 | 143.097 |
| $\mathbf{5}$ | 196.667 | 311.918 | 195.962 | 181.306 | 416.565 | 405.503 | 211.696 | 107.232 | 147.262 |
| $\mathbf{6}$ | 152.185 | 159.319 | 239.577 | 125.594 | 112.734 | 315.512 | 293.860 | 141.199 | 67.536 |
| $\mathbf{7}$ | 89.665 | 99.168 | 90.479 | 129.007 | 60.393 | 64.436 | 160.015 | 145.938 | 72.474 |
| $\mathbf{8}$ | 83.180 | 55.161 | 48.667 | 42.621 | 53.016 | 27.520 | 29.546 | 69.918 | 71.107 |
| $\mathbf{9}$ | 42.756 | 40.942 | 23.412 | 18.776 | 13.988 | 18.103 | 11.127 | 11.256 | 28.445 |
| $\mathbf{1 0}$ | 13.403 | 22.055 | 16.635 | 10.260 | 7.305 | 5.942 | 8.342 | 4.460 | 4.576 |
| $\mathbf{1 1}$ | 17.916 | 6.290 | 8.574 | 7.618 | 4.309 | 3.792 | 2.607 | 2.964 | 2.163 |
| $\mathbf{1 2}$ | 7.275 | 10.241 | 3.388 | 3.754 | 3.545 | 1.790 | 1.912 | 1.400 | 1.663 |
| $\mathbf{1 3}$ | 2.333 | 3.569 | 4.431 | 1.535 | 2.171 | 1.634 | 1.201 | 0.792 | 0.652 |
| $\mathbf{1 4}$ | 1.190 | 1.501 | 1.704 | 1.973 | 0.646 | 1.260 | 1.151 | 0.794 | 0.324 |
| $\mathbf{4 +}$ | $\mathbf{9 1 6 . 7 7}$ | $\mathbf{9 2 7 . 3 2}$ | $\mathbf{8 4 0 . 1 7}$ | $\mathbf{1 0 3 8 . 0 7}$ | $\mathbf{1 1 0 0 . 8 5}$ | $\mathbf{1 1 0 3 . 6 0}$ | $\mathbf{8 3 7 . 3 9}$ | $\mathbf{6 7 6 . 1 5}$ | $\mathbf{5 3 9 . 3 0}$ |


| $\mathbf{Y} / \mathbf{A}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{3}$ | 107.259 | 199.782 | 263.924 | 121.299 | 191.272 | 78.411 | 241.438 | 248.833 | 216.121 |
| $\mathbf{4}$ | 210.333 | 107.762 | 219.043 | 268.317 | 114.076 | 217.942 | 80.614 | 290.122 | 244.433 |
| $\mathbf{5}$ | 190.674 | 181.673 | 117.214 | 230.810 | 248.609 | 115.214 | 196.436 | 80.870 | 247.705 |
| $\mathbf{6}$ | 81.367 | 159.888 | 147.256 | 93.875 | 179.302 | 199.815 | 81.888 | 151.905 | 61.742 |
| $\mathbf{7}$ | 54.429 | 53.756 | 119.857 | 99.867 | 63.752 | 117.147 | 109.910 | 50.932 | 90.326 |
| $\mathbf{8}$ | 15.610 | 26.881 | 30.212 | 74.831 | 63.382 | 33.262 | 57.083 | 52.367 | 25.271 |
| $\mathbf{9}$ | 10.519 | 6.391 | 14.423 | 16.366 | 38.097 | 25.957 | 15.207 | 25.229 | 20.635 |
| $\mathbf{1 0}$ | 10.075 | 4.421 | 3.633 | 7.281 | 8.223 | 15.945 | 10.112 | 6.062 | 9.801 |
| $\mathbf{1 1}$ | 5.699 | 4.351 | 1.694 | 1.477 | 3.250 | 3.220 | 5.906 | 3.081 | 2.131 |
| $\mathbf{1 2}$ | 1.075 | 2.388 | 1.852 | 0.635 | 0.685 | 1.188 | 1.116 | 1.794 | 1.507 |
| $\mathbf{1 3}$ | 0.434 | 0.464 | 0.902 | 0.452 | 0.289 | 0.215 | 0.411 | 0.347 | 0.529 |
| $\mathbf{1 4}$ | 0.476 | 0.230 | 0.097 | 0.142 | 0.110 | 0.044 | 0.042 | 0.073 | 0.089 |
| $\mathbf{4 +}$ | $\mathbf{5 8 0 . 6 9}$ | $\mathbf{5 4 8 . 2 1}$ | $\mathbf{6 5 6 . 1 8}$ | $\mathbf{7 9 4 . 0 5}$ | $\mathbf{7 1 9 . 7 8}$ | $\mathbf{7 2 9 . 9 5}$ | $\mathbf{5 5 8 . 7 3}$ | $\mathbf{6 6 2 . 7 8}$ | $\mathbf{7 0 4 . 1 7}$ |

Table 3.3.20 Cod at Iceland. Division Va. Capelin biomass (' 000 tonnes) used for prediction of cod mean weights-at-age.

| Total <br> Year Biomass |  |
| :---: | ---: |
| 1979 | 3177 |
| 1980 | 2110 |
| 1981 | 1500 |
| 1982 | 1209 |
| 1983 | 2385 |
| 1984 | 3373 |
| 1985 | 3724 |
| 1986 | 4195 |
| 1987 | 3994 |
| 1988 | 3094 |
| 1989 | 2780 |
| 1990 | 2197 |
| 1991 | 2519 |
| 1992 | 3164 |
| 1993 | 3304 |
| 1994 | 3350 |
| 1995 | 4139 |
| 1996 | 5005 |
| 1997 | 4298 |
| 1998 | 3450 |
| 1999 | 3566 |
| 2000 | 3719 |
| 2001 | 3636 |
| 2002 | 3705 |
| 2003 | 3175 |
| Average | 3231 |
|  |  |

Cod at Iceland. Division Va. Landings ('000 tonnes), average fishing mortality of age groups, recruitment (at age 3 in millions), spawning stock at spawning time ('000 tonnes), Harvest Ratio and total biomass ('000 tonnes).

| Year | F5-10 | SSB | Bio4+ | Recruitment | HarvestRatio | Landings |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1955 | 0.31 | 1166 | 2323 | 219 | 0.23 | 545 |
| 1956 | 0.30 | 1094 | 2071 | 301 | 0.24 | 487 |
| 1957 | 0.31 | 1102 | 1934 | 152 | 0.24 | 455 |
| 1958 | 0.31 | 1292 | 1996 | 196 | 0.26 | 517 |
| 1959 | 0.33 | 1071 | 1812 | 133 | 0.25 | 459 |
| 1960 | 0.35 | 912 | 1792 | 172 | 0.26 | 470 |
| 1961 | 0.34 | 749 | 1422 | 277 | 0.23 | 332 |
| 1962 | 0.38 | 789 | 1511 | 247 | 0.26 | 389 |
| 1963 | 0.44 | 673 | 1286 | 270 | 0.32 | 409 |
| 1964 | 0.52 | 594 | 1203 | 312 | 0.36 | 437 |
| 1965 | 0.56 | 461 | 1130 | 171 | 0.34 | 387 |
| 1966 | 0.55 | 425 | 1211 | 254 | 0.29 | 353 |
| 1967 | 0.56 | 501 | 1382 | 186 | 0.24 | 336 |
| 1968 | 0.58 | 582 | 1503 | 186 | 0.25 | 382 |
| 1969 | 0.54 | 669 | 1489 | 139 | 0.27 | 403 |
| 1970 | 0.57 | 662 | 1418 | 283 | 0.33 | 475 |
| 1971 | 0.61 | 495 | 1160 | 177 | 0.39 | 447 |
| 1972 | 0.66 | 421 | 1008 | 261 | 0.39 | 391 |
| 1973 | 0.69 | 431 | 830 | 392 | 0.44 | 369 |
| 1974 | 0.74 | 330 | 916 | 140 | 0.40 | 368 |
| 1975 | 0.76 | 335 | 883 | 224 | 0.41 | 365 |
| 1976 | 0.72 | 287 | 949 | 244 | 0.36 | 346 |
| 1977 | 0.65 | 329 | 1327 | 147 | 0.26 | 340 |
| 1978 | 0.54 | 394 | 1331 | 144 | 0.25 | 330 |
| 1979 | 0.49 | 507 | 1421 | 134 | 0.26 | 366 |
| 1980 | 0.52 | 541 | 1479 | 225 | 0.29 | 432 |
| 1981 | 0.61 | 428 | 1306 | 141 | 0.36 | 465 |
| 1982 | 0.71 | 263 | 965 | 136 | 0.39 | 380 |
| 1983 | 0.72 | 221 | 805 | 343 | 0.37 | 298 |
| 1984 | 0.66 | 229 | 917 | 301 | 0.31 | 282 |
| 1985 | 0.68 | 267 | 927 | 181 | 0.35 | 323 |
| 1986 | 0.77 | 266 | 840 | 86 | 0.43 | 365 |
| 1987 | 0.83 | 255 | 1038 | 129 | 0.38 | 390 |
| 1988 | 0.85 | 201 | 1101 | 108 | 0.34 | 378 |
| 1989 | 0.73 | 286 | 1104 | 166 | 0.33 | 363 |
| 1990 | 0.73 | 336 | 837 | 145 | 0.40 | 335 |
| 1991 | 0.79 | 218 | 676 | 74 | 0.46 | 308 |
| 1992 | 0.84 | 236 | 539 | 148 | 0.49 | 265 |
| 1993 | 0.84 | 220 | 572 | 181 | 0.44 | 251 |
| 1994 | 0.73 | 256 | 581 | 82 | 0.31 | 178 |
| 1995 | 0.53 | 324 | 548 | 156 | 0.31 | 169 |
| 1996 | 0.55 | 270 | 656 | 58 | 0.28 | 181 |
| 1997 | 0.59 | 353 | 794 | 185 | 0.26 | 203 |
| 1998 | 0.67 | 294 | 720 | 166 | 0.34 | 244 |
| 1999 | 0.72 | 323 | 730 | 167 | 0.36 | 265 |
| 2000 | 0.77 | 242 | 559 | 207 | 0.43 | 239 |
| 2001 | 0.78 | 325 | 663 | 69 | 0.35 | 234 |
| 2002 | 0.76 | 357 | 704 | 195 | 0.30 | 208 |
| Mean | 0.61 | 478 | 1133 | 188 | 0.33 | 352 |

Table 3.3.22

| Iceland Cod: VPA and groundfish survey data, N+S |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | :--- |
| 422 2 |  |  |  |  |  |
| 'Yearcl' | 'VPAage3' | Surv4' | 'Surv3' | 'Surv2' | 'Surv1' |
| 1981 | 141 | 4809 | -11 | -11 | -11 |
| 1982 | 136 | 2243 | 3485 | -11 | -11 |
| 1983 | 343 | 8203 | 9556 | 11107 | -11 |
| 1984 | 301 | 10161 | 10310 | 6056 | 1654 |
| 1985 | 181 | 7770 | 7169 | 2886 | 1508 |
| 1986 | 86 | 1407 | 2197 | 736 | 365 |
| 1987 | 129 | 3017 | 2615 | 1645 | 344 |
| 1988 | 108 | 1887 | 1793 | 1179 | 404 |
| 1989 | 166 | 3641 | 3326 | 1627 | 556 |
| 1990 | 145 | 2666 | 3076 | 1713 | 395 |
| 1991 | 74 | 899 | 897 | 482 | 72 |
| 1992 | 148 | 2944 | 2478 | 1501 | 357 |
| 1993 | 181 | 5618 | 4260 | 2903 | 1438 |
| 1994 | 82 | 1606 | 1357 | 548 | 118 |
| 1995 | 156 | 4227 | 2998 | 2239 | 372 |
| 1996 | 58 | 694 | 701 | 556 | 121 |
| 1997 | 185 | 3760 | 5474 | 3298 | 806 |
| 1998 | 166 | 4016 | 3378 | 2790 | 739 |
| 1999 | -11 | 3855 | 4112 | 2172 | 1879 |
| 2000 | -11 | -11 | 4636 | 3807 | 1216 |
| 2001 | -11 | -11 | -11 | 444 | 92 |
| 2002 | -11 | -11 | -11 | -11 | 1117 |

Table 3.3.23

```
Analysis by RCT3 ver3.1 of data from file :
in_2003.dat
Iceland Cod: VPA and groundfish survey data,N+S
Data for 4 surveys over 22 years : 1981-2002
Regression type = C
Tapered time weighting not applied
Survey weighting not applied
Final estimates shrunk towards mean
Minimum S.E. for any survey taken as . 20
Minimum of 3 points used for regression
Forecast/Hindcast variance correction used.
```

Year class $=1998$

| Survey/ <br> Series | Slope | Intercept | Std Error | Rsquare | $\begin{aligned} & \text { No. } \\ & \text { Pts } \end{aligned}$ | Index <br> Value | Predicted Value | Std Error | WAP <br> Weights |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Surv4 | . 65 | -. 23 | . 17 | . 888 | 17 | 8.30 | 5.13 | . 185 | . 268 |
| Surv3 | . 65 | -. 25 | . 15 | . 915 | 16 | 8.13 | 5.03 | . 165 | . 268 |
| Surv2 | . 56 | . 74 | . 12 | . 949 | 15 | 7.93 | 5.22 | . 132 | . 268 |
| Surv1 | . 52 | 1.74 | . 24 | . 783 | 14 | 6.61 | 5.18 | . 273 | . 144 |
|  |  |  |  |  | VPA | Mean = | 4.94 | . 459 | . 051 |

Year class $=1999$

| Survey/ <br> Series | Slope | Intercept | Std Error | Rsquare | No. <br> Pts | Index Value | $\begin{gathered} \text { Predicted } \\ \text { Value } \end{gathered}$ | Std Error | WAP <br> Weights |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Surv4 | . 65 | -. 23 | . 16 | . 889 | 18 | 8.26 | 5.10 | . 178 | . 269 |
| Surv3 | . 65 | -. 27 | . 15 | . 913 | 17 | 8.32 | 5.16 | . 162 | . 269 |
| Surv2 | . 56 | . 76 | . 12 | . 948 | 16 | 7.68 | 5.07 | . 128 | . 269 |
| Surv1 | . 52 | 1.75 | . 23 | . 787 | 15 | 7.54 | 5.66 | . 277 | . 140 |
|  |  |  |  |  | VPA | Mean = | 4.95 | . 447 | . 054 |

Year class $=2000$

| Survey/ <br> Series | Slope | Intercept | Std Error | Rsquare | $\begin{aligned} & \text { No. } \\ & \text { Pts } \end{aligned}$ | Index Value | Predicted Value | Std Error | WAP <br> Weights |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Surv4 |  |  |  |  |  |  |  |  |  |
| Surv3 | . 65 | -. 27 | . 15 | . 913 | 17 | 8.44 | 5.24 | . 163 | . 362 |
| Surv2 | . 56 | . 76 | . 12 | . 948 | 16 | 8.24 | 5.38 | . 130 | . 362 |
| Surv1 | . 52 | 1.75 | . 23 | . 787 | 15 | 7.10 | 5.43 | . 266 | . 204 |
|  |  |  |  |  | VPA | Mean = | 4.95 | . 447 | . 072 |

Year class $=2001$

| Survey/ <br> Series | Slope | Intercept | Std Error | Rsquare | $\begin{aligned} & \text { No. } \\ & \text { Pts } \end{aligned}$ | Index Value | Predicted Value | Std <br> Error | WAP <br> Weights |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Surv4 |  |  |  |  |  |  |  |  |  |
| Surv3 |  |  |  |  |  |  |  |  |  |
| Surv2 | . 56 | . 76 | . 12 | . 948 | 16 | 6.10 | 4.18 | . 137 | . 583 |
| Surv1 | . 52 | 1.75 | . 23 | . 787 | 15 | 4.53 | 4.10 | . 278 | . 301 |
|  |  |  |  |  | VPA | Mean = | 4.95 | . 447 | . 117 |

Table 3.3.23 (Cont'd)


Table 3.3.24

Calculation were done with a spreadsheet: codpr2002.xls
Input data:
Sexual maturity at
spawning time:

| agelyear | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | AV00-02 | Av75-02 | Av85-02 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3 | 0.12 | 0.098 | 0.098 | 0.098 | 0.098 | 0.098 | 0.039 | 0.050 |
| 4 | 0.41 | 0.334 | 0.334 | 0.334 | 0.334 | 0.334 | 0.142 | 0.193 |
| 5 | 0.59 | 0.548 | 0.548 | 0.548 | 0.548 | 0.548 | 0.343 | 0.425 |
| 6 | 0.84 | 0.754 | 0.754 | 0.754 | 0.754 | 0.754 | 0.587 | 0.670 |
| 7 | 0.85 | 0.857 | 0.857 | 0.857 | 0.857 | 0.857 | 0.794 | 0.834 |
| 8 | 0.99 | 0.935 | 0.935 | 0.935 | 0.935 | 0.935 | 0.909 | 0.920 |
| 9 | 0.99 | 0.979 | 0.979 | 0.979 | 0.979 | 0.979 | 0.956 | 0.958 |
| 10 | 0.98 | 0.982 | 0.982 | 0.982 | 0.982 | 0.982 | 0.968 | 0.963 |
| 11 | 1.00 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 0.995 | 0.996 |
| 12 | 1.00 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 0.983 | 0.979 |
| 13 | 1.00 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 0.988 | 0.982 |
| 14 | 1.00 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |

Mean weights in the spawning stock (1/1-31/5 in catches each year)

| agelyear | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{0 0 - 0 2}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3 | 1.025 | 1.050 | 1.050 | 1.050 | 1.050 | 1.050 |
| 4 | 1.498 | 1.607 | 1.607 | 1.607 | 1.607 | 1.607 |
| 5 | 2.159 | 2.391 | 2.385 | 2.385 | 2.385 | 2.347 |
| 6 | 3.236 | 3.532 | 3.477 | 3.473 | 3.473 | 3.381 |
| 7 | 4.655 | 4.955 | 4.852 | 4.810 | 4.806 | 4.830 |
| 8 | 5.957 | 6.397 | 6.382 | 6.296 | 6.259 | 6.428 |
| 9 | 7.881 | 8.048 | 8.048 | 8.048 | 8.048 | 8.048 |
| 10 | 9.458 | 9.327 | 9.327 | 9.327 | 9.327 | 9.327 |
| 11 | 10.231 | 10.035 | 10.035 | 10.035 | 10.035 | 10.035 |
| 12 | 11.736 | 11.873 | 11.873 | 11.873 | 11.873 | 11.873 |
| 13 | 13.172 | 12.660 | 12.660 | 12.660 | 12.660 | 12.660 |
| 14 | 17.442 | 16.070 | 16.070 | 16.070 | 16.070 | 16.070 |

## Mean weights in the catch

| agelyear | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{9 8 - 0 0}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3 | 1.294 | 1.367 | 1.367 | 1.367 | 1.367 | 1.367 |
| 4 | 1.926 | 1.825 | 1.859 | 1.859 | 1.859 | 1.919 |
| 5 | 2.656 | 2.632 | 2.583 | 2.602 | 2.602 | 2.545 |
| 6 | 3.680 | 3.613 | 3.600 | 3.561 | 3.577 | 3.448 |
| 7 | 4.720 | 4.871 | 4.819 | 4.807 | 4.772 | 4.725 |
| 8 | 6.369 | 6.136 | 6.251 | 6.214 | 6.205 | 6.257 |
| 9 | 7.808 | 7.712 | 7.712 | 7.712 | 7.712 | 7.712 |
| 10 | 9.002 | 9.087 | 9.087 | 9.087 | 9.087 | 9.087 |
| 11 | 10.422 | 9.819 | 9.819 | 9.819 | 9.819 | 9.819 |
| 12 | 13.402 | 11.367 | 11.367 | 11.367 | 11.367 | 11.367 |
| 13 | 9.008 | 11.130 | 11.130 | 11.130 | 11.130 | 11.130 |
| 14 | 16.893 | 16.070 | 16.070 | 16.070 | 16.070 | 16.070 |

Table 3.3.24 (Continued)
Selection pattern from AD-CAM:

| agelyear | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{0 0 - 0 2}$ | Used |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3 | 0.044 | 0.042 | 0.070 | 0.085 | 0.089 | 0.058 | 0.078 | 0.078 |
| 4 | 0.246 | 0.217 | 0.258 | 0.250 | 0.280 | 0.234 | 0.255 | 0.255 |
| 5 | 0.446 | 0.546 | 0.561 | 0.568 | 0.514 | 0.471 | 0.518 | 0.518 |
| 6 | 0.748 | 0.782 | 0.962 | 0.850 | 0.829 | 0.803 | 0.828 | 0.828 |
| 7 | 1.033 | 0.988 | 1.000 | 1.125 | 1.019 | 0.988 | 1.044 | 1.044 |
| 8 | 1.144 | 1.139 | 1.070 | 1.113 | 1.177 | 1.156 | 1.149 | 1.149 |
| 9 | 1.220 | 1.242 | 1.167 | 1.123 | 1.192 | 1.255 | 1.190 | 1.190 |
| 10 | 1.409 | 1.302 | 1.240 | 1.220 | 1.268 | 1.327 | 1.271 | 1.271 |
| 11 | 1.499 | 1.346 | 1.251 | 1.189 | 1.204 | 1.237 | 1.210 | 1.637 |
| 12 | 1.607 | 1.435 | 1.335 | 1.256 | 1.242 | 1.261 | 1.253 | 1.637 |
| 13 | 2.655 | 2.349 | 2.178 | 2.036 | 2.022 | 2.072 | 2.043 | 1.637 |
| 14 | 2.655 | 2.349 | 2.178 | 2.036 | 2.022 | 2.072 | 2.043 | 1.637 |
| Ave(5-10) | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |

## Natural Mortality

| agelyear | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 3 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |
| 4 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |
| 5 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |
| 6 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |
| 7 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |
| 8 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |
| 9 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |
| 10 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |
| 11 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |
| 12 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |
| 13 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |
| 14 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |

Table 3.3.24 (Continued)
Mortality proportions
Given
stock before spawning numbers

| agelyear | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{F}$ | $\mathbf{M}$ |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3 | 207.175 | 69.00 | 195.00 | 180.00 | 0.085 | 0.250 |  |
| 4 | 130.817 |  |  |  | 0.180 | 0.250 |  |
| 5 | 87.067 |  |  |  | 0.248 | 0.250 |  |
| 6 | 53.486 |  |  |  | 0.296 | 0.250 |  |
| 7 | 7.477 |  |  |  | 0.382 | 0.250 |  |
| 8 | 7.421 |  |  |  | 0.437 | 0.250 |  |
| 9 | 1.354 |  |  |  | 0.477 | 0.250 |  |
| 10 | 0.839 |  |  |  | 0.250 |  |  |
| 11 | 0.326 |  |  |  | 0.477 | 0.250 |  |
| 12 | 0.065 |  |  |  | 0.477 | 0.250 |  |
| 13 | 0.035 |  |  |  |  | 0.477 | 0.250 |
| 14 | 0.010 |  |  |  |  |  |  |

Table 3.3.24 (Continued)

## Icelandic COD. Division Va.

Prognosis

## Summary

Catch, $\quad$ '000
tonnes

|  | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Opt1 | 265 | 239 | 234 | 208 | 210 | 180 | 180 | 180 |
| Opt2 | 265 | 239 | 234 | 208 | 210 | 210 | 223 | 231 |
| Opt3 | 265 | 239 | 234 | 208 | 210 | 190 | 190 | 190 |
| Opt4 | 265 | 239 | 234 | 208 | 210 | 220 | 220 | 220 |
| Opt5 | 265 | 239 | 234 | 208 | 210 | 250 | 250 | 250 |

## Average fishing mortality of 5-10 years old

|  | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | 2003 | 2004 | 2005 | 2006 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Opt1 | 0.720 | 0.770 | 0.780 | 0.760 | 0.583 | 0.405 | 0.328 | 0.273 |
| Opt2 | 0.720 | 0.770 | 0.780 | 0.760 | 0.583 | 0.485 | 0.444 | 0.414 |
| Opt3 | 0.720 | 0.770 | 0.780 | 0.760 | 0.583 | 0.431 | 0.355 | 0.300 |
| Opt4 | 0.720 | 0.770 | 0.780 | 0.760 | 0.583 | 0.513 | 0.446 | 0.398 |
| Opt5 | 0.720 | 0.770 | 0.780 | 0.760 | 0.583 | 0.600 | 0.554 | 0.526 |

Fishable stock, 4+ in '000 tonnes at the beginnig of the year

|  | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Opt1 | 730 | 559 | 663 | 704 | 766 | 914 | 902 | 1069 |
| Opt2 | 730 | 559 | 663 | 704 | 766 | 914 | 867 | 980 |
| Opt3 | 730 | 559 | 663 | 704 | 766 | 914 | 890 | 1044 |
| Opt4 | 730 | 559 | 663 | 704 | 766 | 914 | 855 | 971 |
| Opt5 | 730 | 559 | 663 | 704 | 766 | 914 | 819 | 897 |

Spawning stock in '000 at the time of spawning

|  | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Opt1 | 323 | 242 | 325 | 357 | 374 | 448 | 531 | 628 |
| Opt2 | 323 | 242 | 325 | 357 | 374 | 440 | 492 | 541 |
| Opt3 | 323 | 242 | 325 | 357 | 374 | 445 | 519 | 605 |
| Opt4 | 323 | 242 | 325 | 357 | 374 | 437 | 484 | 537 |
| Opt5 | 323 | 242 | 325 | 357 | 374 | 428 | 448 | 471 |

Table 3.3.24 (Continued)

## Icelandic COD. Division Va.

## Prognosis - Summary table (nwwg2003)

| 2003 |  |  |  | 2004 |  |  |  | 2005 |  |  |  | 2006 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TAC | $\begin{gathered} \hline 4+ \\ \text { stofn } \\ 4+ \\ \text { stock } \end{gathered}$ | Hr. stofn Sp. stock | $\begin{gathered} F \\ (5-10) \end{gathered}$ | TAC | $\begin{gathered} \hline 4+ \\ \text { stofn } \\ 4+ \\ \text { stock } \end{gathered}$ | Hr. stofn Sp. stock | $\begin{gathered} F \\ (5-10) \end{gathered}$ | TAC | $\begin{gathered} \hline 4+ \\ \text { stofn } \\ 4+ \\ \text { stock } \end{gathered}$ | Hr . stofn Sp. stock | $\begin{gathered} F \\ (5-10) \end{gathered}$ | TAC | $\begin{gathered} \hline 4+ \\ \text { stofn } \\ 4+ \\ \text { stock } \end{gathered}$ | Hr. <br> stofn <br> Sp. <br> stock | $\begin{gathered} F \\ (5-10) \end{gathered}$ |
| 210 | 766 | 374 | 0.583 | 180 | 914 | 448 | 0.405 | 180 | 902 | 531 | 0.328 | 180 | 1069 | 628 | 0.273 |
|  |  |  |  | 210 | 914 | 440 | 0.485 | 223 | 867 | 492 | 0.444 | 231 | 980 | 541 | 0.414 |
|  |  |  |  | 190 | 914 | 445 | 0.431 | 190 | 890 | 519 | 0.355 | 190 | 1044 | 605 | 0.300 |
|  |  |  |  | 220 | 914 | 437 | 0.513 | 220 | 855 | 484 | 0.446 | 220 | 971 | 537 | 0.398 |
|  |  |  |  | 250 | 914 | 428 | 0.600 | 250 | 819 | 448 | 0.554 | 250 | 897 | 471 | 0.526 |

The shaded option corresponds to the harvest control rule.
Table 3.3.25 Cod at Iceland. Division Va. Yield-per-recruit input data

| MFYPR version 1 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Run: final |  |  |  |  |  |  |
| Cod Va (NWWG 2003) |  |  |  |  |  |  |
| Time and date: 11:50 03/05/2003 |  |  |  |  |  |  |
| Fbar age range: 5-10 |  |  |  |  |  |  |
| Age M Mat PF PM SWt Sel CW |  |  |  |  |  |  |
| 3 | 0.20 .044 | 0.085 | 0.25 | 1.3228 | 0.079 | 1.3228 |
| 4 | 0.20 .174 | 0.18 | 0.25 | 1.8442 | 0.314 | 1.8442 |
| 5 | 0.20 .388 | 0.248 | 0.25 | 2.5961 | 0.581 | 2.5961 |
| 6 | 0.20 .621 | 0.296 | 0.25 | 3.5406 | 0.860 | 3.5406 |
| 7 | 0.20 .797 | 0.382 | 0.25 | 4.7361 | 1.052 | 4.7361 |
| 8 | 0.20 .900 | 0.437 | 0.25 | 6.1747 | 1.200 | 6.1747 |
| 9 | 0.20 .948 | 0.477 | 0.25 | 7.6166 | 1.169 | 7.6166 |
| 10 | 0.20 .964 | 0.477 | 0.25 | 9.3120 | 1.138 | 9.3120 |
| 11 | 0.20 .994 | 0.477 | 0.25 | 10.8710 | 1. 051 | 10.8710 |
| 12 | 0.20 .980 | 0.477 | 0.25 | 12.9050 | 1.055 | 12.9050 |
| 13 | 0.20 .984 | 0.477 | 0.25 | 14.2971 | 1.060 | 14.2971 |
| 14 | 0.21 | 0.477 | 0.25 | 15.9137 | 1.050 | 15.9137 |

Weights in kilograms

Table 3.3.26 Cod at Iceland. Division Va. Yield-per-recruit summary.

|  | Fish Mort Yield/R <br> Ages 5-10 | $\mathrm{SSB} / \mathrm{R}$ |  |
| :--- | ---: | ---: | ---: |
| Average Current | 0.76 | 1.67 | 2.20 |
| $\mathbf{F}_{\text {max }}$ | 0.38 | 1.76 | 4.45 |
| $\mathbf{F}_{0.1}$ | 0.20 | 1.63 | 7.68 |
| $\mathbf{F}_{\text {med }}$ | 0.54 | 1.73 | 3.13 |



Figure 3.3.1 Cod at Iceland Division Va. Landings since 1905.


Figure 3.3.2 Landings by gear and year. Upper pictures in tonnes and lower in percentages.


Figure 3.3.3 Cod in division Va. Gillnet landings by mesh size and year.


Figure 3.3.4 Cod in division Va. Catch in numbers by year and age.


Figure 3.3.5 Icelandic cod. Catch curves. Grey lines show $\mathrm{Z}=1$.


Figure 3.3.6 Cod in division Va. Maturity-at-age in the catches.


Figure 3.3.7 Cod in division Va. Mean weight-at-age in the catches.


Figure 3.3.8 Cod in division Va. Mean weight-at-age in the SSB.


Figure 3.3.9.A Cod at Iceland Division Va. Percentages changes in effort for the main geras since 1991.


Figure 3.3.9.B Cod at Iceland Division Va. Percentages changes in cpue for the main gears since 1991.


Figure 3.3.10 Cod in division Va. Total biomass index from the groundfish survey. Index base on the median of all stations where cod was caught is shown for comparison. The scale is 100 thousand tonnes.


Figure 3.3.11 Cod in division Va. Survey indices from the March survey. Numbers by year and age.


Figure 3.3.12 Cod in division Va. Catchcurves from the survey. The grey lines show $\mathrm{Z}=1$


Figure 3.3.13 Cod in division Va. Indices from the groundfish survey vs. index of the same year class in survey a year later.


Figure 3.3.14 Cod in division Va. Survey indices vs. number in stock. Line fitted on original scale


Figure 3.3.15 Cod in division Va. Survey indices vs. number in stock. Line fitted on logscale (power curve)


Figure 3.3.16 Retrospective pattern from assessment runs. The figures show mean fishing mortality of ages 5 to 10.



Figure 3.3.16. (Continued)


Figure 3.3.17 Retrospective patterns from assessment runs. The figures show number of age 4 and older multiplied by the weight in the catches.



Figure 3.3.17. (Continued)


AD-CAM $\ln ($ CNay-observed/CNay-predicted)


AD-CAM $\ln$ (Uay-observed/Uay-predicted)


Figure 3.3.18 Residuals by year and age group from the various models. Solid symbols indicate positive values, open symbols indicate negative values. Bubble area is proportional to magnitude.


TSA - Standardized prediction errors of survey cpue


EX-CAM $\ln ($ CNay-observed/CNay-predicted)

Figure 3.3.18. (Continued)


Coleraine $\ln ($ CNay-observed/CNay-predicted)


Coleraine $\ln$ (Uay-observed/Uay-predicted)

Figure 3.3.18. (Continued)


Figure 3.3.19 Comparison of estimated fishing mortalities in 2002 from different assessment runs.


Figure 3.3.20 Comparison of estimated stock in numbers in 2003 from different assessment runs.


Figure 3.3.21 Estimated 4+ biomass from the various assessment models.


Figure 3.3.22A Yield and fishing mortality


Figure 3.3.22B Spawning stock and recruitment.


Figure 3.3.23A Results of different management options.



Figure 3.3.23B
Estimed age composition of the SSB and the catches in 2004.


Figure 3.3.24 Cod in division Va. Posterior distribution of number in stock at the start of 2003.


Figure 3.3.25 Yield-per-recruit


Figure 3.3.26 Spawning stock biomass and recruitment at age 3


Figure 3.2.27 Cod in division Va. Stock estimate and medium-term prognosis according to the ammended catchrule (errors inmw@age and assessment error included).

### 3.4.1 Introductory comment

Haddock (Melanogrammus aeglefinus) in Icelandic waters is only connected with other haddock stocks due to 0-group and occasionally young fish found in E-Greenland waters originate from the Icelandic stock. The species is distributed all around the Icelandic coast, principally in the relatively warm waters off the west and south coast, in fairly shallow waters ( $50-200 \mathrm{~m}$ depth). Haddock is also found off the North coast and in warm periods a large part of the immature fish can be found in that area. In warm periods the area inhabited by the stock in considerably larger than in cold period. Recent years have been relatively warm and since 1998 recruitment has been exceptionally good with 4 of 5 most recent year classes being strong, something which has not been observed in 40 years. This is probably due to favourable environmental conditions for haddock north of Iceland.

Icelandic haddock was assessed at the North-Western Working Group in 1970 and 1976 but otherwise assessments were conducted by the Marine Research Institute in Iceland until in 1999 when it was again assessed by the NorthWestern Working Group.

### 3.4.2 Trends in landings and fisheries

During the early sixties haddock landings were around 100000 tonnes for five years (Figure 3.4.2.1) After that, landings have been between 40000 and 65000 tonnes. Historically landings by foreign fleets accounted for up to half of the total landed catch. Since 1976 fisheries by foreign nations have been negligible except a small catch by the Faroese. Haddock landings are subject to fluctuations, reflecting variability in stock biomass and recruitment, which is very variable.

The landings in 2002 are estimated as 50400 tonnes, increasing from 39600 tonnes in 2001. In last year the forecasted landings for the year 2002 were 45000 tonnes.

In 2002, $61 \%$ of landings were by demersal trawl, $7 \%$ by Danish seine, $28 \%$ by long line and $4 \%$ by gillnets. The share of bottom trawl increased from 2001 but the share of longline decreased, but is still high as it has been since 1998.

### 3.4.3 Catch-at-age

Catch-at-age for 2002 for the Icelandic fishery is provided in Table 3.4.3.1. Catch-at-age is calculated by 3 fleets and two time intervals. The time intervals are January-May and June-December and the fleets are gillnets, long line and bottom trawl. Hand lines are included with the long line fleet. Danish seine (as well as minor units such as pelagic trawl and other gears which are dragged or hauled) are included in the trawl feet. The Faroese catch that is caught by long line is included in that category. Numbers sampled in 2002 are given in the table below.

| Gear | Season | Region | No of <br> length <br> samples | Number <br> length <br> measured | Number of <br> age samples | Number. <br> aged | Number <br> weighed |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Bottom trawl | Jan-May | South | 138 | 10897 | 20 | 1041 | 839 |
| Bottom trawl | Jan-May | North | 46 | 687 | 3 | 149 | 149 |
| Bottom trawl | June-Dec | South | 99 | 7041 | 19 | 927 | 786 |
| Bottom trawl | June-Dec | North | 150 | 10644 | 13 | 599 | 498 |
| Danish seine | Jan-May | South | 211 | 614 | 5 | 249 | 249 |
| Danish seine | Jan-May | North | 29 | 348 | 2 | 100 | 50 |
| Danish seine | June-Dec | South | 48 | 801 | 6 | 294 | 294 |
| Gillnet | Jan-May | South | 106 | 2416 | 8 | 392 | 391 |
| Gillnet | Jan-May | North | 14 | 394 | 0 | 0 | 0 |
| Gillnet | June-Dec | South | 138 | 1136 | 7 | 350 | 350 |
| Gillnet | June-Dec | North | 63 | 3581 | 0 | 0 | 0 |
| Longline | Jan-May | South | 487 | 28876 | 36 | 1818 | 1240 |
| Longline | Jan-May | North | 620 | 13999 | 17 | 840 | 495 |
| Longline | June-Dec | South | 312 | 13650 | 31 | 1639 | 1286 |
| Longline | June-Dec | North | 641 | 26244 | 39 | 1959 | 987 |

For comparison the calculations of catch in numbers by age were done by 3 gears, 2 regions (North and South) and 2 time intervals giving identical results.

Catch-at-age in 2002 deviated more from last years predictions than is usual for this stock. As may be seen in the table below the deviations are not large except for the year class 1995 (age 7) where the proportion caught exceeded predictions by $75 \%$.

| Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Forecast \% | 6.1 | 33.5 | 38.3 | 13.8 | 2.1 | 4.8 | 0.5 | 0.4 |
| Catch \% | 2.7 | 28.2 | 43 | 13.6 | 3.0 | 8.3 | 0.7 | 0.5 |

Figure 3.4.3.1 shows the catch in number plotted on log scale. The curves indicate that total mortality was high or close to 1 for the oldest haddock but is possibly decreasing in the most recent years. The 1976 year class is shown for comparison but the fishing mortality was low around 1980. Figure 3.4.3.1 indicates that CV in these data is low. Shephard Nicholson model gives a CV of $23 \%$ for age groups 3-8.

### 3.4.4 Weight and maturity-at-age

Mean weight-at-age in the catch is shown in table 3.4.4.1.
Mean weight-at-age in the stock for 1982-2002 is given in Table 3.4.4.2. Those data are calculated from the Icelandic groundfish survey. Weights for 1985-1992 were calculated using a length-weight relationship which is the mean from the years 1993-2003. Weights from 1993 onwards are based on weighting of fish in the groundfish survey each year. Stock weights prior to 1985 have been taken to be the mean of 1985-2002 weights.

Both stock and catch weights have been relatively low since 1990 compared to the eighties. Since 1990 the weights have not shown any apparent trend but it seems like the large year classes (1990 and 1995) grow slower. Four of the most recent year classes are large so drop in mean weights would not be unexpected in coming years. No drop is seen in the catch weights from 2002 but the survey weights drop in 2003 for age groups 3 to 5 which are the most abundant age groups.

Maturity-at-age data are given in table 3.4.4.3. They show high maturity-at-age in recent years compared to earlier years and maturity-at-age increased from 2002 to 2003. Maturity-at-age data from 1985 onwards are taken from the groundfish survey but maturity-at-age in catches January - May is used 1980 to 1984.

### 3.4.5 Survey and cpue data

Haddock is one of the most abundant fishes in the Icelandic groundfish survey in March, being caught in large number-at-age 1 and becoming fully recruited at age 2 or 3. Age disaggregated indices from the March survey are given in table 3.4.5.11 and indices from the fall survey in table 3.4.5.2.

The index of total biomass from the Icelandic groundfish survey in March is shown in Figure 3.4.5.1. It was at record low in 2000 but has increased since then due to good recruitment and is in the year 2003 the highest since the series started. The increase between 2002 and 2003 is much larger than predicted and the catch curves and Hjörleifsson Pálsson plot (Figures 3.4.5.4 and 3.4.5.3) suggest that the 2003 survey is an outlier for year classes 1998 and 1999. The median index shown in figure 3.4.5.1, calculated as the number of stations where haddock is found, times the median of the haddock catch at those stations shows similar increase as the traditional Cochran index showing that the increase is not only at the stations with most haddock.

Figure 3.4.5.2 shows the total index from the autumn survey. The autumn survey does not decrease as much in 2000 and shows less increase in the most recent years.

In figure 3.4.5.5 survey indices from the March survey are plotted against VPA estimates with regression lines based on all data until 1999 and $r^{2}$ in the fit of these lines included. The figure shows that the survey indices are good measure of stock size and the relationship between survey index and number in stock is close to linear for all agegroups. The most recent estimates are shown as intersection of dashed.

The survey indicates that in some of recent years unusually high proportion of the incoming year classes have been in
the northern part of the survey area (figure 3.4.5.6) in areas where fishing effort has been relatively light. As described in WD\#34 the relatively little overlap of the recruiting year classes and the fishery in recent years can explain why discards have reduced in recent years and recent year classes have progressively become stronger in every new survey.

CPUE from the commercial fleet is shown in figure 3.4.5.7. The figure is calculated from records where more than $50 \%$ of the catch is haddock and shows increase from 2001 to 2002, especially for danish seine and bottom trawl. Effort by the most important fleets, fishing haddock decreased between 2001 and 2002 (figure 3.4.5.8).

### 3.4.6 Stock Assessment

Last years assessment was based on XSA, using groundfish survey indices for age 2-9 for tuning but in addition a number of other models and tuning data were explored. This year the same procedure was used i.e to examine many different models and tuning data. To have more than 1 person running the assessments was also considered important and therefore 4 scientists made an assessment of Icelandic haddock this year.

### 3.4.6.1 Tuning and estimation of fishing mortality

The SPALY (Same Procedure As Last Year) XSA run used survey indices from age 1985-2003, ages 2-9 for tuning. Shrinkage was set to 2 years and 2 ages with $\mathrm{SE}=0.5$ Catchability of all age groups was independent of stock size. In addition to the SPALY a number of other models and settings were used. The results of some of the runs are shown in the table below.

|  | F4-7 2002 | Biomass <br> $3+$ | Std. Err in <br> $3+$ bio | N8-2003 | N4-2003 | N3-2003 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Spaly XSA survey 2-9 | 0.61 | 209 |  | 1.8 | 93 | 150 |
| XSA No shrinkage survey 2-9 | 0.55 | 235 |  | 2 | 104 | 190 |
| XSA Bottom trawl CPUE+ <br> Age 2 from March survey | 0.511 | 182 |  | 3.3 | 68 | 153 |
| XSA Survey from age 1 | 0.61 | 192 |  | 1.9 | 82 | 132 |
| TSA | 0.62 | 197 | 14 | 2.0 | 73 | 146 |
| Excam | 0.45 | 258 |  | 2 | 113 | 173 |
| Adapt like program | 0.6 | 201 |  | 1.7 | 88 | 134 |
| ADCAM | 0.63 | 192 | 21 | 1.6 | 84 | 127 |
| ADCAM EXCAM weights | 0.55 | 235 |  | 2 | 106 | 164 |
| ADCAM Variable M | 0.61 | 225 | 25 | 1.6 | 98 | 161 |
| ADCAM Autumn survey | 0.66 | 141 | 22 | 1.5 | 57 | 70 |
| Autumn survey Variable M | 0.64 | 178 |  | 1.5 | 69 | 117 |

The EXCAM is a is a statistical catch-at-age model implemented in Excel (Described in working paper 32)
The ADCAM is a statistical catch-at-age model written in AD-model builder (described in working paper 33 last year) was used. The settings used for haddock were;

- Fishing mortality was estimated for every year and age.
- Recruitment was assumed to be lognormally distributed around a fixed mean with the CV of the lognormal distribution estimated. This term can be looked at as the P-shrinkage in the model. The estimate was of the CV was 0.8 to be compared with estimated CV of the survey shown in figure 3.4.6.2.
- CV of commercial catch data and of survey indices as function of age are estimated. The CV of the commercial catch is a parabola but estimated separately for each age in the survey (change from last year when it was also a $2^{\text {nd }}$ order polynomial (Figure 3.4.6.2). Correlation of residuals of different age groups in the survey was estimated as a $1^{\text {st }}$ order AR model.
- Catchability in survey was independent of stock size for all ages.
- Fishing mortality of each age group was random walk with standard deviation specified as proportion of the estimated CV in the catch-at-age data. In the input file the process error (variability in F ) is specified to be larger than the measurement error for the younger ages but the measurement error is specified to be larger for the older age groups. This could be looked at as an F shrinkage term in the model.
- The model estimates standard deviation on survey and age disaggregated catches. The division of the standard deviation in catches between process (random walk of $F$ ) and measurement error can not be estimated.

A Cohort model where the survivors were estimated and the stock back calculated by Popes equation was also used. In this model (referred to as ADAPT like) in the table above the parts involving the survey, SSB-recruitment and prognosis were the same as in the catch-at-age model but the model does not include any kind of F-shrinkage term.

In addition to the standard model different alternatives were tested, checking the effect of different shrinkage terms, weighting of survey age groups and hidden mortality.

The results presented in the table vary considerably with biomass ( $3+$ ) ranging from 190 to 260 thousand tonnes in 2003 if the March survey is used for tuning but 140-180 if the autumn survey is used for tuning. For comparison the 2003 March survey indicates on its own that the biomass (3+) in 2003 exceeds 300 kT . Figure 3.4.6.4 show the residuals from the ADCAM run using the March survey (giving 190 kT ) and figure 3.4.6.5 observed vs. predicted survey biomass from the same run. Both figures show the large positive residuals in the 2003 survey and that the model does not follow the data. The same positive residuals in 2003 can be seen in the SPALY-XSA output (table 3.4.6.2) as well as in most of the other runs Included with the TSA results from Guðmundur Guðmundsson was a sentence that the "results from the 2003 survey were outliers, specially for the 1998 and 1999 year classes and should be treated with care". Figure 3.4.6.5 indicates that the models are not following the 2003 survey very closely.

As expected the results from the model are not driven by the name of the model but the assumptions behind the model. Therefore it should not be surprising that results of the ADCAM results using as far as possible the same assumptions as in the EXCAM model give similar results. What is most important here is the relative weight on different age groups in the survey. All the year classes 1998 - 2000 have progressively become stronger in the survey so reducing weight of younger ages in the survey is going to give higher estimate of these year classes. The relative weights used in the EXCAM were estimated last year assuming that they followed a $2^{\text {nd }}$ degree polynomial in age while this year the relative weights are estimated for each age group in the ADCAM model. The same feature can be seen by comparing the stock estimates from XSA using ages 2-9 for tuning and XSA using ages 1-9 for tuning, including the indices at age 1 leads to lower stock estimates. Whether it is more appropriate to estimate the weights on different age groups independently or assume a $2^{\text {nd }}$ order polynomial is not obvious. Estimating the weight on different age groups independently can in some cases lead to shift in the weights if a new year with contradicting data is added, possibly introducing things like can happen in multifleet XSA runs.

Inclusions of terms similar to P and F shrinkage is questionable for stock like the Icelandic haddock where reasonable survey indices are available from age 1 . As the situation is now with good recruitment and possibly reducing F , the effect of shrinkage terms is to reduce estimated stock sizes.

The ADCAM runs where the hidden mortality of the youngest age groups is a function of the overlap between the fisheries (WD\#34) and the spatial distribution from the surveys, explains part of the discrepancy seen in the surveys in recent years when large part of the recruiting year classes have been observed in areas with relatively little fishing effort. The model does though only explain the discrepancy from age 1 to 3 but the discrepancy between the years 2002 and 2003 for age year classes 1998 and 1999 (figure 3.4.5.3) are not explained by this model. The model findings that "hidden mortality" was much reduced in recent years is supported by discard data (table ???) so there are indications that hidden mortality caused by the fisheries is important for this stock.

Using the autumn survey for tuning gives lower stock estimates and higher fishing mortality than using the March survey. Figures 3.4.6.6 and 3.4.4.7 do on the other hand indicate show positive residuals in the 2002 survey indicating an underestimate of the stock. (Direct estimate by the 2002 autumn survey gives B3+ $\approx 165 \mathrm{kT}$ at the start of 2003)

The standard error of the Biomass (3+) by some of the models is given in the table above. It is smaller than the largest difference between different models and it is clearly an underestimate of the "real uncertainty" in the assessment. The presence of the "outliers" from 2003 is not reflected in the standard errors but a use of a distribution with heavier tails than the normal distribution would probably help in getting more realistic confidence intervals.

So, how to select the "final run"? All of the presented runs seem equally plausible and there is no way to point to any result being the most likely one. In last year the SPALY XSA run was used but RTC3 results were used for the youngest age groups. This year results from the AD-model builder program ADCAM were selected for prognosis. As seen in the table above the results given by that model is on the conservative side regarding results using the March survey but higher than estimates using CPUE from the trawler fleet or the autumn survey. The main reason for selecting this model for point estimates is "the convenience" of having the recruitment, assessment and prognosis in the same package allowing to carry the uncertainity in assessment and recruitment into the prognosis. Also the procedure of
selecting some age groups out of one model and other age groups out of another model is questionable as the estimates of different age groups within a model are correlated. It is though again emphasised that all the models presented here were used to select a point estimator.

### 3.4.7 Recruitment estimates

The discussion about recruitment estimates has been treated in last section under stock assessment. Many of the models used for stock assessment also give recruitment estimates but in addition to those model 2 recruitment models, RTC3 and a time-series model from Guðmundur Guðmundsson were used. The table below shows results of the various models. The results of different models differ considerably, especially for year class 1999 but estimates of this year class in the survey vary considerably from 1 survey to another. This year class does on the other hand hardly belong to the group "recruitment" although it is included in the table.

| Recruitment (million 2 year old.) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year class | RTC3 | EXCAM | XSA <br> Age 1 | TSA REC | ADCAM | ADCAM <br> reduced P- <br> shrink | Survey <br> 2002 | Survey <br> 2003 | Last years <br> estimate |
| 1998 | 83 |  |  |  |  |  | 116 | 166 |  |
| 1999 | 113 | 188 | 138 | 120 | 141 | 152 | 134 | 234 | 112 |
| 2000 | 164 | 213 | 162 | 158 | 156 | 173 | 199 | 187 | 155 |
| 2001 | 48 | 55 | 53 | 55 | 40 | 43 |  | 78 | 29 |
| 2002 | 123 | 223 | 148 | 144 | 113 | 130 |  | 170 | 170 |

Input to the RTC3 model is shown in table 3.4.7.1 and output from the model in table 3.4.7.2

### 3.4.8 Prediction of catch and biomass

### 3.4.8.1 Input data

The input data for the prediction are shown in Table 3.4.8.1

For the short-term catch prediction and stock biomass calculations, the mean weight-at-age 3-8 in the catches in 2003 were predicted using regression analysis, where the mean weight-at-age was predicted by the mean weight of the same year class in the previous year. The procedure is then repeated between the years 2003 and 2004.

For the stock weights survey weights for the year 2003 were used for that year but for the year 2004 mean weight-at-age was predicted from the mean weight of the same year class in the survey in 2003

The exploitation pattern was taken as the mean exploitation pattern from 1998-2002.
Maturity is taken to be the mean of the 2001-2003 values.

Stock in number in the year 2003 and recruitment in 2004 - 2005 was obtained from the ADCAM model and the prognosis were done by the ADCAM model and an Excel spreadsheet that has been used at the MRI in Reykjavík for a number of years.

A TAC constraint of 65000 tonnes was applied to the prediction for the year 2003. The estimate was the sum of the TAC for the fishing year starting September $1^{\text {st }} 2002$ that was remaining in the beginning of 2003 and $27 \%$ of the estimated TAC for the fishing year 2003-2004 but 27\% of the TAC for the fishing year 2002-2003 was taken in the year 2002. In the prognosis last year a TAC constraint of 45000 was used for the year 2002 while the estimated landings now were 50000 tonnes. The picture now is that $\mathrm{F}_{4-7}$ did not change from 2001 to 2002 so status quo F would have been a better assumption? But the problem is not so simple as last year $\mathrm{F}_{4-7}$ in 2001 was considered to be 0.75 compared to 0.63 this year so does $\mathbf{F}_{\mathrm{sq}}$ help. The year before the prognosis of TAC in the assessment year was an underestimate. Figure 3.4.6.3 shows retros of in predictions 4 years $\mathrm{F}_{4-7}$ ahead using the observed catches and they do not indicate a big problem if "the landings are correctly estimated". For information $\mathbf{F}_{\text {sq }}$ will lead to catches of 80000 t in 2003

The short-term prognosis were done using the ADCAM model and also the MFDP program at Ices. In the ADCAM prognosis the proposed $\mathbf{F}_{\mathrm{pa}}$ of 0.47 was used for the years 2004 and 2005. Assessment error was assumed to be
lognormal with $15 \%$ CV and no autocorrelation. Variations in stock and catch weights were assumed to be lognormal with $13 \% \mathrm{CV}$ and an autocorrelation of 0.35 between years The same deviations in weights were applied to all age groups the same year. Errors in weight-at-age and assessment errors were not correlated which they probably should be. The ADCAM model was also used for deterministic prognosis to see if the results were identical to MFDP.

For the long-term yield and spawning stock biomass per recruit, the exploitation pattern was taken as the mean relative fishing mortality from 1980-2001. Mean weight-at-age in the stock and the maturity ogive are means from 1980-2001. Mean weight-at-age in the catch is the mean from 1980-2001. Input data for long-term yield-per-recruit are given in Table 3.4.8.2.

### 3.4.8.2 Biological reference points

The yield-per-recruit is shown in table 3.4.8.3. and figure 3.4.8.1

Compared to the estimated fishing mortality of $\mathrm{F}_{4-7}=0.63$ for $2002, \mathbf{F}_{\max }=0.45$ and $\mathbf{F}_{0.1}=0.18$.

Yield-per-recruit at $\mathbf{F}_{\text {max }}$ corresponds to 0.88 kg . (Table 3.4.8.3). Mean weights as in the most recent years would give lower yield-per-recruit.

A plot of spawning stock biomass and recruitment from 1981-2000 is shown in Figure 3.4.8.2 and a plot of recruitment vs. spawning stock in figure 3.4.8.3.

In the year 2000 the working group proposed provisional $\mathbf{F}_{\mathrm{pa}}$ set to the $\mathbf{F}_{\text {med }}$ value of 0.47 and since no further work has been done since then on reference points for this stock that value will be used for $\mathbf{F}_{\mathrm{pa}}$ this year. Since $1986 \mathrm{~F}_{4-7}$ has exceeded $\mathbf{F}_{\max }$ and for only 4 years since 1960 has $\mathrm{F}_{4-7}$ been lower than $\mathbf{F}_{\mathrm{pa}}$.

The SGPRP proposed $\mathbf{B}_{\text {loss }}$ as candidate for $\mathbf{B}_{\mathrm{pa}}$ at its meeting in February 2003. The working group did not discuss this matter further.

TAC for Icelandic fish stock is given for fishery years which are from September $1^{\text {st }}$. each year to August $3^{\text {rd }}$ the following year. $1 / 3^{\text {rd }}$ of the fishing year 2003/2004 falls within the calendar year 2003 and $2 / 3^{\text {rd }}$ within the calendar year 2004. The TAC for the next fishing year will therefore be $1 / 3^{\text {rd }}$ of the landings in 2003 plus $2 / 3^{\text {rd }}$ of the advice for 2004.

### 3.4.8.3 Projection of catch and biomass

At the beginning of 2003, the biomass of age $3+$ is predicted to be 190000 t with a spawning stock of 129000 t . (Tables 3.4.8.4)

With a catch of 65000 t in 2003, fishing mortality is estimated to be 0.48 , the biomass of age $3+$ is predicted to be 205 000 t in the beginning of the year 2003 and the spawning stock biomass 147000 t

The predictions indicate that the annual catches could be around or above 80000 tonnes for at least the next 2-4 years if the $\mathrm{F}=0.47$ will be used as HCR. $\mathrm{F}=0.47$ leads to 80000 for the calendar year 2004 (Fishing year??)

Figures 3.4.8.3 and 3.4.8.4 show the output of the short-term prognosis including errors in mean weight-at-age and assessment errors.

### 3.4.9 Management considerations

For more than a decade fishing mortality on haddock has been high with $\mathrm{F}_{4-7}$ between 0.6 and 0.8 since 1986 . The advice for 2003 was based on $\mathrm{F}_{\text {med }}$.

The short-term predictions do not show much advantage in reducing fishing mortality. It must though be born in mind that a number of factors, like discard, hidden mortality due to mesh penetration and reduction of mean weight-at-age by removal of the largest individuals of each age group are not included in these predictions.

As described in working paper 34 and in section 3.4..?? discard and other hidden mortality, most likely caused by the fisheries might be a potential problem for this stock. The model described predicts that hidden mortality by mobile
fishing gear might be important for age 1 and 2 haddock and if it proceeds being low, catches exceeding 100000 would not be unrealistic in coming years. An important management consideration is how to reduce this possibly hidden mortality. Distribution of recruits is something out of our control but the amount and distribution of the fishing effort can be controlled. The first option would be to reduce the total effort towards haddock and hence the fishing fleet would be able to catch their allowable quota outside the areas where the smallest haddock might be found. If the premises of the prognosis hold, fishing effort will already be reduced considerably in 2003. Another way could be area closure to protect the recruits. Part of the problem might come from fisheries which are not targeting haddock, both the nephrops fishery and demersal fisheries where cod, saithe or flatfishes are the main target species. In those fisheries the same applies, effort reduction should reduce this problem.

The model runs including variable M predict catches around 100 kT in 2004 assuming $\mathrm{F}=0.47$. If the model predictions of hidden mortality caused by the fisheries are correct "optimal" F is most likely lower than 0.47 so the advice of 80 kT based on the base run and $\mathrm{F}=0.47$ should not be changed.

### 3.4.10 Comments on the assessment

The current assessment was done using only groundfish survey indices for tuning. .
Fishing mortality on haddock increased after 1985 (Figure 3.4.6.2.) The high fishing mortality was at least partly due to an overestimation of the stock biomass through the use of catch weights that are $20-25 \%$ higher than survey weights which have been used in the assessment since 1999.

The assessment presented here gives $\mathrm{F}_{4-7}=0.63$ in 2002 which is the same as the estimate for 2001. Last year the estimate for $\mathrm{F}_{4-7}$ in 2001 was 0.75 .

This years assessment gives a more optimistic view of the stock than last years assessment with the changes driven by the results of the 2003 March survey. The biomass of age 3+ at the start of 2003 is now predicted to be 190000 t (SSB 129000 t ) but was predicted to be $168000 \mathrm{t}(\mathrm{SSB} 98000 \mathrm{t})$ in last years assessment. This is in spite of higher than predicted catches in 2002 ( 50 kT vs. 45 kT )

In this years assessment a number of different models were used and the range of results investigated. The point estimates selected for prognosis come from a model called ADCAM. Those estimates are close to estimates done by many other models (XSA, TSA) but the results from the most recent survey indicate that the stock in larger as do some of the assessment models that were explored.

Table 3.4.2.1 Haddock in Division Va Landings by nation
Table 1.1. Icelandic haddock. Landings by nation.

| Country | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 1010 | 1144 | 673 | 377 | 268 | 359 | 391 | 257 |
| Faroe Islands | 2161 | 2029 | 1839 | 1982 | 1783 | 707 | 987 | 1289 |
| Iceland | 52152 | 47916 | 61033 | 67038 | 63889 | 47216 | 49553 | 47317 |
| Norway €UK | 11 | 23 | 15 | 28 | 3 | 3 | + |  |
| Total | 55334 | 51112 | 63560 | 69425 | 65943 | 48285 | 50933 | 48863 |
| HADDOCK Va |  |  |  |  |  |  |  |  |
| Country | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 |
| Belgium | 238 | 352 | 483 | 595 | 485 | 361 | 458 | 248 |
| Faroe Islands | 1043 | 797 | 606 | 603 | 773 | 757 | 754 | 911 |
| Iceland | 39479 | 53085 | 61792 | 66004 | 53516 | 46098 | 46932 | 58408 |
| Norway UK | 1 | + |  |  |  |  |  | 1 |
| Total | 40761 | 54234 | 62881 | 67202 | 53774 | 47216 | 48144 | 59567 |
| HADDOCK Va |  |  |  |  |  |  |  |  |
| Country | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Belgium |  |  |  |  |  |  |  |  |
| Faroe Islands | 758 | 664 | 340 | 639 | 624 | 968 | 609 | 878 |
| Iceland | 60061 | 56223 | 43245 | 40795 | 44557 | 41199 | 39038 | 49591 |
| Norway UK | + | 4 |  |  |  |  |  |  |
| Total | 60819 | 56891 | 43585 | 41434 | 45481 | 42167 | 39647 | 50469 |

Table 3.4.3.1 Haddock in division Va. Catch in number by year and age.

| year/age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1981 | 0 | 0.5 | 4.9 | 17 | 6 | 2.8 | 1.81 | 0.17 |
| 1982 | 0 | 0.3 | 2.7 | 10.7 | 14.1 | 2.3 | 1.17 | 0.82 |
| 1983 | 0 | 0.7 | 1.5 | 4.6 | 10.3 | 8.8 | 0.87 | 0.24 |
| 1984 | 0.1 | 0.8 | 5 | 1.2 | 4.9 | 3.8 | 4.45 | 0.17 |
| 1985 | 0.4 | 1.8 | 5 | 6.1 | 0.8 | 1.6 | 2.48 | 2.21 |
| 1986 | 0.2 | 3.7 | 3.8 | 4.9 | 5.8 | 0.5 | 0.85 | 0.9 |
| 1987 | 2.2 | 7.6 | 7.5 | 2.7 | 2.2 | 1.2 | 0.15 | 0.21 |
| 1988 | 0.1 | 10.1 | 15.9 | 5.6 | 1.3 | 1 | 0.58 | 0.06 |
| 1989 | 0.1 | 2.6 | 23.1 | 9 | 9.7 | 3.1 | 0.5 | 0.51 |
| 1990 | 0.4 | 2.6 | 8 | 23.8 | 6.7 | 0.9 | 0.17 | 0.14 |
| 1991 | 2.5 | 1.3 | 3.9 | 6.7 | 13.6 | 3 | 0.4 | 0.05 |
| 1992 | 2.7 | 7.3 | 4.2 | 4.2 | 4 | 5.9 | 1.31 | 0.13 |
| 1993 | 0.2 | 11.6 | 12.6 | 3.2 | 1.8 | 1.5 | 2.26 | 0.38 |
| 1994 | 0.3 | 3 | 27 | 10.7 | 1.6 | 0.8 | 0.4 | 0.7 |
| 1995 | 2.4 | 6.3 | 5.7 | 23.4 | 5.6 | 0.6 | 0.26 | 0.21 |
| 1996 | 1.5 | 9 | 7.1 | 4.8 | 14 | 2.4 | 0.23 | 0.09 |
| 1997 | 1.4 | 3.7 | 11.1 | 4.9 | 2.5 | 5 | 0.69 | 0.05 |
| 1998 | 0.2 | 8.1 | 6 | 8.4 | 2.4 | 1.5 | 1.88 | 0.21 |
| 1999 | 1.1 | 1.5 | 16.9 | 4.8 | 5 | 0.9 | 0.59 | 0.51 |
| 2000 | 2.4 | 6.5 | 2.3 | 13.8 | 2.1 | 1.8 | 0.36 | 0.2 |
| 2001 | 2.2 | 11.3 | 7.1 | 1.5 | 6.2 | 0.7 | 0.48 | 0.1 |
| 2002 | 1 | 10.6 | 16.2 | 5.1 | 1.1 | 3.1 | 0.24 | 0.18 |

Table 3.4.4.1 Haddock in division Va Weight-at-age in the catches.

| Year/ <br> age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1982 | 330 | 819 | 1365 | 1649 | 2329 | 3012 | 3384 | 3965 |
| 1983 | 655 | 958 | 1436 | 1827 | 2355 | 2834 | 3569 | 4308 |
| 1984 | 980 | 1041 | 1476 | 2105 | 2460 | 3028 | 3014 | 3807 |
| 1985 | 599 | 1002 | 1783 | 2201 | 2727 | 3431 | 3783 | 4070 |
| 1986 | 867 | 1187 | 1755 | 2377 | 2710 | 3591 | 3760 | 4135 |
| 1987 | 446 | 1048 | 1629 | 2373 | 2984 | 3550 | 4483 | 4667 |
| 1988 | 468 | 808 | 1474 | 2230 | 2934 | 3545 | 3769 | 4574 |
| 1989 | 745 | 856 | 1170 | 2010 | 2879 | 4109 | 4035 | 4706 |
| 1990 | 357 | 716 | 1039 | 1542 | 2403 | 3458 | 4186 | 4969 |
| 1991 | 409 | 868 | 1111 | 1546 | 2035 | 2849 | 3464 | 4642 |
| 1992 | 320 | 856 | 1253 | 1597 | 2088 | 2529 | 3133 | 4022 |
| 1993 | 420 | 756 | 1372 | 1870 | 2360 | 2888 | 2975 | 3442 |
| 1994 | 568 | 720 | 1058 | 1742 | 2380 | 2785 | 3447 | 3156 |
| 1995 | 457 | 874 | 1145 | 1366 | 2079 | 2853 | 3251 | 3899 |
| 1996 | 387 | 841 | 1189 | 1528 | 1816 | 2641 | 3499 | 3526 |
| 1997 | 450 | 829 | 1192 | 1663 | 1934 | 2360 | 3059 | 3010 |
| 1998 | 689 | 777 | 1166 | 1692 | 2312 | 2379 | 2882 | 3417 |
| 1999 | 616 | 866 | 1096 | 1638 | 2205 | 2681 | 2863 | 3229 |
| 2000 | 518 | 951 | 1314 | 1461 | 2096 | 2679 | 3181 | 3438 |
| 2001 | 542 | 933 | 1451 | 1759 | 1836 | 2309 | 2966 | 3123 |
| 2002 | 573 | 918 | 1256 | 1741 | 2192 | 2224 | 2844 | 3392 |

Table 3.4.4.2 Haddock in division Va Weight-at-age in the stock

| Year/ <br> age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1985 | 35 | 244 | 567 | 1187 | 1673 | 2372 | 2768 | 3199 | 3334 | 3718 |
| 1986 | 35 | 239 | 671 | 1134 | 1944 | 2400 | 3192 | 3295 | 3731 | 3675 |
| 1987 | 31 | 162 | 550 | 1216 | 1825 | 2605 | 3031 | 3644 | 3838 | 4099 |
| 1988 | 37 | 176 | 456 | 974 | 1831 | 2697 | 3104 | 3483 | 3321 | 4357 |
| 1989 | 26 | 182 | 440 | 886 | 1510 | 2382 | 3011 | 3502 | 3198 | 3681 |
| 1990 | 29 | 184 | 456 | 839 | 1234 | 1966 | 2677 | 3055 | 3269 |  |
| 1991 | 31 | 176 | 500 | 1002 | 1406 | 1885 | 2498 | 3757 | 3656 | 5458 |
| 1992 | 28 | 157 | 503 | 894 | 1365 | 1892 | 2326 | 2938 | 3684 | 5120 |
| 1993 | 41 | 169 | 384 | 879 | 1487 | 1766 | 2548 | 2538 | 3227 |  |
| 1994 | 33 | 179 | 401 | 696 | 1242 | 1683 | 1641 | 2693 | 1991 |  |
| 1995 | 37 | 164 | 444 | 763 | 1071 | 1856 | 2667 | 5312 | 1313 |  |
| 1996 | 41 | 174 | 447 | 806 | 1072 | 1474 | 2160 | 2407 | 4803 | 2186 |
| 1997 | 50 | 173 | 423 | 818 | 1224 | 1426 | 1917 | 2397 | 3694 | 3573 |
| 1998 | 41 | 202 | 404 | 742 | 1232 | 1738 | 2015 | 2333 | 3081 |  |
| 1999 | 34 | 205 | 479 | 719 | 1198 | 1967 | 2381 | 2798 | 2929 | 5313 |
| 2000 | 29 | 179 | 552 | 888 | 1167 | 1777 | 2620 | 2924 | 3155 | 3668 |
| 2001 | 36 | 188 | 487 | 1052 | 1433 | 1502 | 2165 | 2758 |  | 3900 |
| 2002 | 63 | 172 | 474 | 891 | 1465 | 1955 | 2143 | 1998 | 3662 | 4981 |
| 2003 | 40 | 230 | 412 | 801 | 1268 | 1873 | 3139 | 2343 | 3301 | 4191 |

Table 3.4.4.3 Haddock in division Va Sexual maturity-at-age in the stock and the survey.

| Year/ <br> age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1985 | 0 | 1.6 | 14.4 | 53.6 | 57.8 | 76.5 | 76.6 | 96.1 | 93.4 |
| 1986 | 0 | 2.1 | 20.5 | 41.3 | 67.3 | 84.5 | 88.4 | 95.2 | 98.6 |
| 1987 | 0 | 2.2 | 13.7 | 42.6 | 53.5 | 77.8 | 77.6 | 100 | 96.9 |
| 1988 | 0 | 1.3 | 22.1 | 39.4 | 76.7 | 79.4 | 92.8 | 91.4 | 100 |
| 1989 | 0 | 4.1 | 20.2 | 53.2 | 72.7 | 81.8 | 99.8 | 100 | 100 |
| 1990 | 0 | 11.4 | 33.4 | 63.4 | 81.5 | 84.3 | 91.8 | 88.2 | 100 |
| 1991 | 0 | 6.3 | 22.4 | 59.3 | 73.9 | 81.7 | 89.4 | 49.5 | 100 |
| 1992 | 0 | 5 | 22.7 | 42 | 79.9 | 90.1 | 90.1 | 85.8 | 100 |
| 1993 | 0.5 | 12.4 | 36.4 | 48.8 | 67.4 | 90.6 | 97.7 | 91 | 86.8 |
| 1994 | 3.5 | 25.6 | 31.7 | 59.9 | 78.5 | 85.9 | 100 | 87.8 | 100 |
| 1995 | 0 | 12.9 | 48 | 39.2 | 75.3 | 75.4 | 61.3 | 98.5 | 100 |
| 1996 | 0 | 19.8 | 37.9 | 59.7 | 65.1 | 78.8 | 74 | 94.7 | 89.7 |
| 1997 | 1.5 | 9.3 | 43.4 | 58.4 | 68.2 | 75 | 78.4 | 87.9 | 100 |
| 1998 | 0 | 3.1 | 48.5 | 68 | 77.5 | 73.6 | 85.2 | 89.9 | 100 |
| 1999 | 0 | 5 | 39.5 | 67.9 | 72.3 | 75 | 89.6 | 76.3 | 92 |
| 2000 | 0 | 10.6 | 25.6 | 62.7 | 80.5 | 86.7 | 87.3 | 100 | 77.7 |
| 2001 | 0.2 | 10 | 37.8 | 52 | 75.2 | 89.7 | 92.1 | 91.7 |  |
| 2002 | 0 | 4.7 | 28.4 | 63 | 80 | 93.5 | 92.8 | 100 | 100 |
| 2003 | 0.5 | 6.2 | 34.7 | 68.5 | 86.7 | 92.2 | 94.6 | 100 | 100 |

Table 3.4.5.1 Icelandic haddock. Age disaggregated survey indices from the groundfish survey in March

| Year/ age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 28.15 | 32.72 | 18.34 | 23.65 | 26.54 | 3.73 | 10.98 | 4.88 | 5.64 | 0.51 |
| 1986 | 123.95 | 108.51 | 59.07 | 12.8 | 16.38 | 13.2 | 0.98 | 2.77 | 1.26 | 2.32 |
| 1987 | 22.22 | 296.28 | 163.63 | 57.08 | 13.17 | 11.17 | 8.09 | 0.58 | 1.28 | 0.84 |
| 1988 | 15.77 | 40.71 | 184.77 | 88.86 | 22.86 | 1.36 | 2.25 | 1.87 | 0.18 | 0.28 |
| 1989 | 10.58 | 23.35 | 41.53 | 146.71 | 44.9 | 12.74 | 0.85 | 0.84 | 0.41 | 0.28 |
| 1990 | 70.48 | 31.86 | 27.25 | 39.06 | 91.79 | 30.87 | 3.44 | 0.9 | 0.23 | 0 |
| 1991 | 89.73 | 145.95 | 41.55 | 17.83 | 20.27 | 32.55 | 7.67 | 0.3 | 0.1 | 0.11 |
| 1992 | 18.15 | 211.43 | 138.4 | 35.54 | 16.56 | 13.14 | 15.93 | 2.21 | 0.18 | 0.07 |
| 1993 | 29.99 | 37.65 | 245.06 | 87.3 | 11.15 | 3.86 | 1.66 | 4.46 | 0.88 | 0 |
| 1994 | 58.54 | 61.34 | 39.83 | 142.62 | 42.41 | 6.93 | 2.89 | 1.42 | 4.07 | 0 |
| 1995 | 35.89 | 82.53 | 48.09 | 19.74 | 68.41 | 7.66 | 1.31 | 0.11 | 0.34 | 0 |
| 1996 | 95.25 | 66.3 | 121 | 36.93 | 19.11 | 39.77 | 5.84 | 0.62 | 0.13 | 0.12 |
| 1997 | 8.57 | 119.13 | 50.88 | 52.99 | 10.86 | 7.28 | 10.58 | 1.37 | 0.06 | 0.03 |
| 1998 | 23.12 | 18.07 | 108.27 | 28.25 | 23.32 | 4.64 | 3.47 | 4.57 | 0.33 | 0 |
| 1999 | 80.73 | 86.21 | 25.8 | 98.18 | 12.9 | 9.6 | 1.42 | 1.7 | 1.03 | 0.03 |
| 2000 | 60.58 | 90.44 | 45.03 | 8.54 | 24.63 | 2.94 | 1.62 | 0.41 | 0.15 | 0.45 |
| 2001 | 81.33 | 148.06 | 115.04 | 22.16 | 4.09 | 10.56 | 0.93 | 0.57 | 0 | 0.1 |
| 2002 | 21.14 | 298.28 | 201 | 112.78 | 23.25 | 3.52 | 7 | 0.31 | 0.34 | 0.11 |
| 2003 | 111.96 | 97.85 | 282.83 | 244.83 | 112.28 | 18.05 | 2.58 | 4.43 | 0.48 | 0.85 |

Table 3.4.5.2 Icelandic haddock. Age disaggregated survey indices from the groundfish survey in October.

| Year/a <br> ge |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1995 | 93.95 | 162.64 | 184.92 | 51.4 | 24.27 | 42.47 | 5.74 | 0.56 | 0 | 0.07 | 0 |
| 1996 | 12.45 | 347.52 | 93.69 | 77.33 | 16.52 | 6.35 | 15.27 | 1.28 | 0 | 0 | 0 |
| 1997 | 49.84 | 29.63 | 200.21 | 59.25 | 39.34 | 7.12 | 5.79 | 6.35 | 0.29 | 0 | 0 |
| 1998 | 183.18 | 79.7 | 33.41 | 138.33 | 19.47 | 13.6 | 4.52 | 4.36 | 1.68 | 0 | 0 |
| 1999 | 204.63 | 343.81 | 57.78 | 26.55 | 96.25 | 10.51 | 8.97 | 0.45 | 1.49 | 0.31 | 0 |
| 2000 | 56.59 | 157.27 | 240.32 | 41.42 | 7.05 | 26.77 | 1.8 | 2.73 | 0.07 | 0.21 | 0.28 |
| 2001 | 50.18 | 331.24 | 253.85 | 155.73 | 31.35 | 3.53 | 12.14 | 0.64 | 0.95 | 0 | 0.2 |
| 2002 | 137.95 | 76.53 | 213.48 | 171.33 | 84.46 | 16.88 | 2.49 | 2.14 | 0.85 | 0.09 | 0 |

Table 3.4.6.1 Haddock in division Va. Input data for tuning.
Groundfish survey age 31985 - 2002

| 1985 | 32.7 |
| :--- | ---: |
| 1986 | 108.5 |
| 1987 | 296.3 |
| 1988 | 40.7 |
| 1989 | 23.4 |
| 1990 | 31.9 |
| 1991 | 146.0 |
| 1992 | 212.3 |
| 1993 | 37.2 |
| 1994 | 61.2 |
| 1995 | 83.2 |
| 1996 | 71.3 |
| 1997 | 120.4 |
| 1998 | 18.2 |
| 1999 | 86.5 |
| 2000 | 91.0 |
| 2001 | 148.1 |
| 2002 | 298.3 |

Groundfish survey age 2-8 shifted.

| 1984 | 18.3 | 23.7 | 26.5 | 3.7 | 11.0 | 4.9 | 5.6 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1985 | 59.1 | 12.8 | 16.4 | 13.2 | 1.0 | 2.8 | 1.3 |
| 1986 | 163.6 | 57.1 | 13.2 | 11.2 | 8.1 | 0.6 | 1.3 |
| 1987 | 184.8 | 88.9 | 22.9 | 1.4 | 2.3 | 1.9 | 0.2 |
| 1988 | 41.5 | 146.7 | 44.9 | 12.7 | 0.9 | 0.8 | 0.4 |
| 1989 | 27.3 | 39.1 | 91.8 | 30.9 | 3.4 | 0.9 | 0.2 |
| 1990 | 41.6 | 17.8 | 20.3 | 32.6 | 7.7 | 0.3 | 0.1 |
| 1991 | 138.4 | 35.5 | 16.6 | 13.2 | 15.9 | 2.2 | 0.2 |
| 1992 | 245.1 | 87.3 | 11.2 | 3.9 | 1.7 | 4.5 | 0.9 |
| 1993 | 39.8 | 142.6 | 42.4 | 6.9 | 2.9 | 1.4 | 4.1 |
| 1994 | 48.1 | 19.7 | 68.4 | 7.7 | 1.3 | 0.1 | 0.3 |
| 1995 | 121.0 | 36.9 | 19.1 | 39.8 | 5.8 | 0.6 | 0.1 |
| 1996 | 50.9 | 53.0 | 10.9 | 7.3 | 10.6 | 1.4 | 0.1 |
| 1997 | 108.3 | 28.3 | 23.3 | 4.6 | 3.5 | 4.6 | 0.3 |
| 1998 | 25.8 | 98.2 | 12.9 | 9.6 | 1.4 | 1.7 | 1.0 |
| 1999 | 45.0 | 8.5 | 24.6 | 2.9 | 1.6 | 0.4 | 0.2 |
| 2000 | 115.0 | 22.2 | 4.1 | 10.6 | 0.9 | 0.6 | 0.0 |
| 2001 | 201.0 | 112.8 | 23.3 | 3.5 | 7.0 | 0.3 | 0.3 |
| 2002 | 282.8 | 244.8 | 112.3 | 18.1 | 2.6 | 4.4 | 0.5 |

Table 3.4.6.2 Haddock Va. Output from XSA.
Lowestoft VPA Version 3.1
24/04/2003 12:08


Time-series weights :
Tapered time weighting not applied

Catchability analysis :
Catchability independent of stock size for all ages
Catchability independent of age for ages >= 7

Terminal population estimation :
Survivor estimates shrunk towards the mean $F$
of the final 2 years or the 2 oldest ages.
S.E. of the mean to which the estimates are shrunk $=0.500$

Minimum standard error for population
estimates derived from each fleet $=0.300$

Prior weighting not applied

Tuning converged after 19 iterations

```
Regression weights
    ,1.000,1.000, 1.000, 1.000, 1.000, 1.000, 1.000, 1.000, 1.000, 1.000
Fishing mortalities
    Age, 1993, 1994, 1995, 1996, 1997, 1998, 1999, 2000, 2001, 2002
        2,0.006,0.008, 0.038,0.047,0.015,0.016, 0.025,0.022,0.016,0.006
        3, 0.098, 0.116, 0.234, 0.196, 0.161, 0.119, 0.148, 0.205, 0.141, 0.098
        4, 0.369, 0.348, 0.331, 0.446, 0.397, 0.426, 0.389, 0.375, 0.363, 0.308
        5, 0.657, 0.619, 0.578, 0.513, 0.642, 0.596, 0.744, 0.644, 0.441, 0.485
        6, 0.768, 0.810, 0.793, 0.848, 0.576, 0.787, 0.895, 0.848, 0.686, 0.712
        7,0.928,0.910, 0.916, 1.036, 0.872, 0.828, 0.843, 1.007, 0.809, 0.932
        8,0.852,0.697, 0.997, 1.152, 0.989, 1.029, 0.958, 0.979, 0.853, 0.764
        9,0.874,0.709, 1.022, 1.173, 0.924, 0.959, 0.913, 1.072, 0.868, 0.903
1
XSA population numbers (Thousands)
```

|  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | , | 2, | 3 , | 4, | 5, | 6, | 7 , | 8, |


| 993 | 3.75E+04, | 1.37E+05, | $4.53 \mathrm{E}+04$ | 7.27E+03 | 3.68E+03, | 2.75E+03, | 4.36E+03, | $7.19 E+02$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 4.13E+04, | $3.05 \mathrm{E}+04$, | 1.02E+05, | $2.57 \mathrm{E}+04$, | $3.09 \mathrm{E}+03$, | 1.40E+03, | $8.89 \mathrm{E}+02$, | 1.52E+03 |
| 1995 | $7.07 \mathrm{E}+04$, | 3.35E+04, | $2.22 \mathrm{E}+04$ | $5.88 \mathrm{E}+04$, | 1.13E+04, | 1.12E+03, | $4.61 \mathrm{E}+02$, | $3.63 \mathrm{E}+02$ |
| 1996 | $3.50 \mathrm{E}+04$, | $5.57 \mathrm{E}+04$ | $2.17 \mathrm{E}+04$ | $1.31 \mathrm{E}+04$ | $2.70 \mathrm{E}+04$ | $4.19 \mathrm{E}+03$, | 3.68E+02, | 1.39E+02 |
| 1997 | 9.89E+04, | 2.74E+04, | $3.75 \mathrm{E}+04$ | 1.14E+04, | $6.41 \mathrm{E}+03$, | 9.46E+03, | 1.22E+03, | $9.53 \mathrm{E}+01$ |
| 1998 | 1.45E+04, | $7.97 \mathrm{E}+04$ | $1.91 \mathrm{E}+04$ | 2.06E+04, | 4.91E+03, | 2.95E+03, | $3.24 \mathrm{E}+03$, | $3.71 \mathrm{E}+02$ |
| 1999 | 4.86E+04, | 1.17E+04, | $5.80 \mathrm{E}+04$ | 1.02E+04, | 9.31E+03, | 1.83E+03, | 1.05E+03, | $9.49 \mathrm{E}+02$ |
| 2000 | 1.18E+05, | $3.88 \mathrm{E}+04$, | 8.25E+03, | 3.22E+04, | 3.97E+03, | 3.12E+03, | $6.44 \mathrm{E}+02$, | $3.31 \mathrm{E}+02$ |
| 2001 | 1.55E+05, | 9.49E+04, | $2.59 \mathrm{E}+04$ | 4.64E+03, | 1.38E+04, | 1.39E+03, | 9.32E+02, | 1.98E+02 |
| 2002 | 1.84E+05, | 1.25E+05, | $6.75 \mathrm{E}+04$, | 1.47E+04, | $2.44 \mathrm{E}+03$, | 5.70E+03, | 5.07E+02, | $3.25 E+02$ |

## Table 3.4.6.2 (Cont'd)

Estimated population abundance at 1st Jan 2003
$, \quad 0.00 \mathrm{E}+00,1.50 \mathrm{E}+05,9.28 \mathrm{E}+04,4.06 \mathrm{E}+04,7.43 \mathrm{E}+03,9.82 \mathrm{E}+02,1.84 \mathrm{E}+03,1.93 \mathrm{E}+02$

Taper weighted geometric mean of the VPA populations:

$$
, \quad 5.62 \mathrm{E}+04,3.94 \mathrm{E}+04,2.68 \mathrm{E}+04,1.54 \mathrm{E}+04,7.86 \mathrm{E}+03,3.22 \mathrm{E}+03,1.15 \mathrm{E}+03,3.93 \mathrm{E}+02
$$

Standard error of the weighted Log(VPA populations) :

$$
, \quad 0.7595, \quad 0.7961, \quad 0.7463, \quad 0.7588, \quad 0.8737, \quad 0.8754, \quad 0.9454, \quad 1.0346
$$

Log catchability residuals.

```
Fleet : 2ara CPU
    Age , 1984, 1985, 1986, 1987, 1988, 1989, 1990, 1991, 1992
    2 , 99.99, -0.46, -0.02, 0.35, -0.37, -0.35, 0.14, 0.39, 0.01
    3,No data for this fleet at this age
    4, No data for this fleet at this age
    5 , No data for this fleet at this age
    6, No data for this fleet at this age
    7 , No data for this fleet at this age
    8, No data for this fleet at this age
Age , 1993, 1994, 1995, 1996, 1997, 1998, 1999, 2000, 2001, 2002
    2,-0.22, 0.18,-0.05, 0.50, -0.02, 0.01, 0.37, -0.48, -0.26, 0.27
    3,No data for this fleet at this age
    4 , No data for this fleet at this age
    5, No data for this fleet at this age
    6 , No data for this fleet at this age
    7, No data for this fleet at this age
    8, No data for this fleet at this age
```

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age, | 2 |
| :---: | :---: |
| Mean Log q, | -4.3463, |
| S.E (Log q), | 0.3073, |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.
Age, Slope, t-value, Intercept, RSquare, No Pts, Reg s.e, Mean Q
2, 0.97,
0.256 ,
4.52,
0.87 ,
18, $0.31,-4.35$,

Fleet : SUR CPU

| Age | , | 1984, | 1985, | 1986, | 1987, | 1988, | 1989, | 1990, | 1991, | 1992 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | , | -0.35, | 0.09, | 0.34 , | -0.16, | -0.40, | -0.24, | 0.38 , | 0.31, | 0.12 |  |
| 3 |  | -0.20, | -0.30, | 0.46, | 0.13, | -0.03, | -0.11, | -0.25, | 0.58, | 0.26 |  |
| 4 | , | 0.04 , | 0.01, | 0.42 , | 0.21, | 0.10, | 0.04 , | -0.16, | 0.26, | 0.09 |  |
| 5 | , | 0.34, | 0.08, | 0.56, | -0.87, | 0.54, | 0.47 , | -0.16, | 0.27 , | -0.20 |  |
| 6 |  | 0.89 , | -0.13, | 0.96 , | -0.10, | -0.31, | 0.38 , | 0.00, | 0.06 , | -0.71 |  |
| 7 | , | -0.06, | 0.51, | 0.59 , | 0.61 , | -0.03, | 0.96 , | -1.01, | -0.17, | -0.08 |  |
| 8 | , | 0.09, | -0.31, | 0.92, | 0.52, | 0.43 , | 0.23, | -0.11, | -0.30, | 0.11 |  |
| Age | , | 1993, | 1994, | 1995, | 1996, | 1997, | 1998, | 1999, | 2000, | 2001, | 2002 |
| 2 |  | -0.20, | -0.11, | 0.31, | 0.15 , | -0.16, | 0.33 , | -0.32, | -0.27, | 0.01 , | 0.17 |
| 3 |  | -0.05, | -0.51, | 0.14 , | -0.05, | 0.00 , | 0.14 , | -0.36, | -0.55, | 0.12, | 0.58 |
| 4 |  | 0.16 , | -0.19, | 0.04, | -0.39, | -0.22, | -0.11, | -0.61, | -0.47, | 0.11 , | 0.67 |
| 5 |  | 0.54, | -0.65, | 0.13, | -0.13, | -0.33, | -0.23, | -0.57, | -0.53, | 0.10, | 0.63 |
| 6 |  | 0.50, | -0.09, | 0.09, | -0.12, | -0.06, | -0.50, | -0.90, | -0.67, | -0.03, | 0.74 |
| 7 |  | 0.34, | -1.64, | 0.37 , | 0.02 , | 0.24, | 0.36, | -0.59, | -0.55, | -0.64, | 0.76 |
| 8 |  | 0.88, | -0.30, | -0.45, | -0.07, | -0.33, | -0.06, | -0.62, | 99.99, | -0.19, | 0.84 |

## Table 3.4.6.2 (Cont'd)

Mean log catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age , | 2, | 3, | 4, | 5, | 6, | 7, | 8 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log q, | -4.1397, | -4.2151, | -4.2642, | -4.3486, | -4.3769, | -4.4967, | -4.4967, |
| S.E (Log q), | 0.2636, | 0.3265, | 0.3049, | 0.4579, | 0.5186, | 0.6557, | 0.4768, |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.
Age, Slope , t-value , Intercept, RSquare, No Pts, Reg s.e, Mean Q

| 2, | 0.99, | 0.100, | 4.20, | 0.90, | 19, | 0.27, | -4.14, |
| ---: | ---: | ---: | ---: | ---: | :--- | :--- | :--- |
| 3, | 0.87, | 1.512, | 5.06, | 0.89, | 19, | 0.27, | -4.22, |
| 4, | 0.99, | 0.100, | 4.33, | 0.84, | 19, | 0.31, | -4.26, |
| 5, | 1.05, | -0.342, | 4.08, | 0.72, | 19, | 0.49, | -4.35, |
| 6, | 0.99, | 0.089, | 4.44, | 0.70, | 19, | 0.53, | -4.38, |
| 7, | 0.94, | 0.329, | 4.71, | 0.62, | 19, | 0.63, | -4.50, |
| 8, | 0.97, | 0.269, | 4.50, | 0.83, | 18, | 0.47, | -4.42, |

1

Terminal year survivor and $F$ summaries :
Age 2 Catchability constant w.r.t. time and dependent on age

```
Year class = 2000
```

| Fleet, |  | Estimated, Survivors, | Int, s.e, | $\begin{aligned} & \text { Ext, } \\ & \text { s.e, } \end{aligned}$ | Var, <br> Ratio, | N, | Scaled, Weights, | $\begin{aligned} & \text { Estimated } \\ & \mathrm{F} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2ara CPU | , | 195660., | 0.316 , | 0.000, | 0.00, | 1 | 0.399 , | 0.005 |
| SUR CPU | , | 177365., | 0.300 , | 0.000, | 0.00, | 1, | 0.442, | 0.005 |
| F shrinkage mean |  | 47951., | 0.50, |  |  |  | 0.160, | 0.019 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | ---: | ---: | ---: | ---: | ---: |
| at end of year, | s.e, | s.e, | Ratio, | Ration |  |
| $149626 .$, | 0.20, | 0.38, | 3, | 1.927, | 0.006 |

Age 3 Catchability constant w.r.t. time and dependent on age

Year class $=1999$

| Fleet, |  | Estimated, | Int, | Ext, | Var, | N, | Scaled, | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| , |  | Survivors, | s.e, | s.e, | Ratio, | , | Weights, | F |
| 2 ara CPU | , | 71527., | 0.316 , | 0.000, | 0.00, | 1, | 0.289, | 0.126 |
| SUR CPU | , | 120844., | 0.223 , | 0.284, | 1.27, | 2, | 0.581 , | 0.076 |
| F shrinkage mean |  | 50595., | 0.50, |  |  |  | 0.129, | 0.174 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | ---: | ---: | ---: | ---: | ---: |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| $92781 .$, | 0.17, | 0.23, | 4, | 1.358, | 0.098 |

1
Age 4 Catchability constant w.r.t. time and dependent on age
Year class $=1998$

| Fleet, |  | Estimated, Survivors, | Int, s.e, | $\begin{aligned} & \text { Ext, } \\ & \text { s.e, } \end{aligned}$ | $\begin{gathered} \text { Var, } \\ \text { Ratio, } \end{gathered}$ |  | Scaled, Weights, | $\begin{gathered} \text { Estimated } \\ \mathrm{F} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 ara CPU | , | 25206., | 0.316, | 0.000, | 0.00, | 1, | 0.206, | 0.458 |
| SUR CPU | , | 49151. | 0.182 , | 0.286 , | 1.57, | 3 , | 0.662, | 0.261 |
| F shrinkage mean |  | 32566. | 0.50, |  |  |  | 0.132 , | 0.371 |

Weighted prediction :

## Table 3.4.6.2 (Cont'd)

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | ---: | ---: | ---: | ---: | :---: |
| at end of year, | s.e, | s.e, | Ratio, |  |  |
| $40576 .$, | 0.15, | 0.22, | 5, | 1.419, | 0.308 |

Age 5 Catchability constant w.r.t. time and dependent on age
Year class $=1997$


Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :--- | ---: | ---: | ---: | ---: | ---: |
| at end of year, | s.e, | S.e, | , | Ratio, |  |
| $7425 .$, | 0.16, | 0.17, | 6, | 1.085, | 0.485 |

1
Age 6 Catchability constant w.r.t. time and dependent on age

Year class = 1996

| Fleet, |  | Estimated, | Int, | Ext, | Var, | N, | Scaled, | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| , |  | Survivors, | s.e, | s.e, | Ratio, | , | Weights, | F |
| 2ara CPU | , | 996., | 0.316, | 0.000 , | 0.00, | 1, | 0.124, | 0.705 |
| SUR CPU | , | 1032., | 0.173 , | 0.226 , | 1.31, | 5, | 0.609, | 0.687 |
| F shrinkage mean | , | 870., | 0.50, |  |  |  | 0.268, | 0.775 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | ---: | ---: | ---: | ---: | ---: |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| $982 .$, | 0.17, | 0.15, | 7, | 0.849, | 0.712 |

Age 7 Catchability constant w.r.t. time and dependent on age

Year class $=1995$

| Fleet, |  | Estimated, | Int, | Ext, | Var, | N, | Scaled, | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| , |  | Survivors, | s.e, | s.e, | Ratio, | , | Weights, | F |
| 2ara CPU | , | 1806., | 0.316, | 0.000, | 0.00, | 1, | 0.072, | 0.942 |
| SUR CPU | , | 1800., | 0.203, | 0.215, | 1.06, | 6, | 0.465, | 0.944 |
| F shrinkage mean |  | 1880., | 0.50, |  |  |  | 0.464, | 0.918 |

Weighted prediction :

| Survivors, | Int, | Ext, | N, | Var, | F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| at end of year, | s.e, | s.e, | , | Ratio, |  |
| $1837 .$, | 0.25, | 0.12, | 8, | 0.496, | 0.932 |

1
Age 8 Catchability constant w.r.t. time and age (fixed at the value for age) 7

Year class $=1994$


## Table 3.4.6.2 (Cont'd)

```
Weighted prediction :
Survivors, Int, Ext, N, Var, F
at end of year, s.e, s.e, g' Ratio, 
Age 9 Catchability constant w.r.t. time and age (fixed at the value for age) }
Year class = 1993
\begin{tabular}{|c|c|c|c|c|c|c|c|c|}
\hline Fleet, & & Estimated, & Int, & Ext, & Var, & & Scaled, & Estimated \\
\hline & & Survivors, & s.e, & s.e, & Ratio, & & Weights, & F \\
\hline 2ara CPU & , & 103., & 0.316, & 0.000, & 0.00 , & & 0.014, & 0.931 \\
\hline SUR CPU & , & 84., & 0.294, & 0.104, & 0.36, & 7, & 0.237, & 1.058 \\
\hline F shrinkage mean & & 117., & 0.50, & & & & 0.749, & 0.857 \\
\hline
\end{tabular}
Weighted prediction :
Survivors, Int, Ext, N, Var, F
at end of year, s.e, s.e, , Ratio
        108., 0.38, 0.11, 9, 0.279, 0.903
```

Table 3.4.6.2 (Cont'd)

Run title : Icelandic Haddock. Run 3.
At 24/04/2003 12:08
Terminal Fs derived using XSA (With F shrinkage)
Table 8 Fishing mortality (F) at age

| YEAR | 1982 | 1983 | 1984 | 1985 |
| :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |
| 2 | 0.0013 | 0.0000 | 0.0033 | 0.0114 |
| 3 | 0.0404 | 0.0228 | 0.0344 | 0.1284 |
| 4 | 0.1339 | 0.3063 | 0.2217 | 0.3310 |
| 5 | 0.3459 | 0.3584 | 0.4211 | 0.4613 |
| 6 | 0.4489 | 0.6653 | 0.8044 | 0.6073 |
| 7 | 0.8749 | 0.5654 | 0.5493 | 0.6617 |
| 8 | 1.1687 | 1.0573 | 0.6323 | 0.8836 |
| 9 | 1.0336 | 0.8197 | 0.5959 | 0.7676 |
| FBAR4-7 | 0.4509 | 0.4738 | 0.4991 | 0.5154 |

Table 8 Fishing mortality (F) at age

| YEAR | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| AGE |  |  |  |  |  |  |  |  |  |
| 2 | 0.0024 | 0.0148 | 0.0031 | 0.0032 | 0.0223 | 0.0345 | 0.0178 | 0.0064 | 0.0075 |
| 3 | 0.1282 | 0.1218 | 0.0857 | 0.0769 | 0.1418 | 0.0824 | 0.1368 | 0.0984 | 0.1163 |
| 4 | 0.4473 | 0.4163 | 0.4054 | 0.2884 | 0.3563 | 0.3310 | 0.4190 | 0.3686 | 0.3478 |
| 5 | 0.6437 | 0.6655 | 0.6362 | 0.4652 | 0.5467 | 0.5777 | 0.7040 | 0.6568 | 0.6193 |
| 6 | 1.1422 | 0.6995 | 0.7759 | 0.9288 | 0.6856 | 0.7124 | 0.8387 | 0.7681 | 0.8099 |
| 7 | 0.9190 | 0.7767 | 0.8095 | 0.9543 | 0.7218 | 0.7645 | 0.8026 | 0.9283 | 0.9101 |
| 8 | 0.9787 | 0.8294 | 1.1814 | 1.4489 | 0.9214 | 0.9169 | 0.9761 | 0.8518 | 0.6972 |
| 9 | 0.9919 | 0.6837 | 0.9316 | 1.1667 | 0.8150 | 0.8565 | 0.9372 | 0.8739 | 0.7087 |
| FBAR4- <br> 7 | 0.7880 | 0.6395 | 0.6568 | 0.6592 | 0.5776 | 0.5964 | 0.6911 | 0.6804 | 0.6718 |
|  |  |  |  |  |  |  |  |  |  |

Run title : Icelandic Haddock. Run 3.
At 24/04/2003 12:08
Terminal Fs derived using XSA (With F shrinkage)

| YEAR | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | $\begin{aligned} & \text { FBAR* } \\ & { }_{*-* *} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |
| 2 | 0.0375 | 0.0474 | 0.0155 | 0.0159 | 0.0248 | 0.0222 | 0.0159 | 0.0061 | 0.0147 |
| 3 | 0.2338 | 0.1961 | 0.1614 | 0.1192 | 0.1481 | 0.2047 | 0.1411 | 0.0984 | 0.1481 |
| 4 | 0.3308 | 0.4460 | 0.3972 | 0.4259 | 0.3890 | 0.3753 | 0.3628 | 0.3083 | 0.3488 |
| 5 | 0.5783 | 0.5134 | 0.6422 | 0.5963 | 0.7444 | 0.6441 | 0.4411 | 0.4854 | 0.5235 |
| 6 | 0.7931 | 0.8480 | 0.5764 | 0.7875 | 0.8948 | 0.8482 | 0.6863 | 0.7119 | 0.7488 |
| 7 | 0.9156 | 1.0358 | 0.8716 | 0.8282 | 0.8428 | 1.0068 | 0.8094 | 0.9320 | 0.9160 |
| 8 | 0.9965 | 1.1521 | 0.9889 | 1.0285 | 0.9581 | 0.9792 | 0.8531 | 0.7645 | 0.8656 |
| 9 | 1.0219 | 1.1726 | 0.9239 | 0.9593 | 0.9133 | 1.0715 | 0.8677 | 0.9034 | 0.9475 |
| $\begin{aligned} & \hline \text { 0FBAR } \\ & 4-7 \\ & \hline \end{aligned}$ | 0.6545 | 0.7108 | 0.6219 | 0.6595 | 0.7178 | 0.7186 | 0.5749 | 0.6094 |  |

Table 3.4.6.2 (Cont'd)

Run title : Icelandic Haddock. Run 3.
At 24/04/2003 12:08
Terminal Fs derived using XSA (With F shrinkage)
Table 10 Stock number-at-age (start of year) numbers*10**-3

| YEAR | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| AGE |  |  |  |  |  |  |
| 2 | 42211 | 30160 | 19932 | 41756 | 89202 | 167933 |
|  |  |  |  |  |  |  |
| 3 | 7980 | 34515 | 24692 | 16265 | 33800 | 72855 |
| 4 | 23796 | 6275 | 27620 | 19533 | 11712 | 24343 |
| 5 | 40447 | 17041 | 3782 | 18117 | 11485 | 6131 |
| 6 | 43132 | 23431 | 9749 | 2032 | 9351 | 4940 |
| 7 | 4336 | 22541 | 9863 | 3571 | 906 | 2443 |
| 8 | 1871 | 1480 | 10486 | 4662 | 1509 | 296 |
| 9 | 1400 | 476 | 421 | 4562 | 1578 | 464 |
| TOTAL | 165174 | 135920 | 106545 | 110498 | 159543 | 279404 |


| YEAR | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |
| 2 | 47637 | 26666 | 22362 | 80244 | 170353 | 37502 | 41275 | 70693 | 35049 |
| 3 | 135467 | 38882 | 21762 | 17905 | 63471 | 137007 | 30506 | 33540 | 55746 |
| 4 | 52809 | 101801 | 29478 | 15462 | 13499 | 45322 | 101660 | 22235 | 21735 |
| 5 | 13144 | 28825 | 62467 | 16902 | 9092 | 7269 | 25667 | 58779 | 13077 |
| 6 | 2580 | 5696 | 14820 | 29606 | 7765 | 3682 | 3086 | 11313 | 26990 |
| 7 | 2010 | 972 | 1842 | 6113 | 11888 | 2748 | 1398 | 1124 | 4191 |
| 8 | 920 | 732 | 307 | 733 | 2330 | 4362 | 889 | 461 | 368 |
| 9 | 106 | 231 | 141 | 100 | 240 | 719 | 1524 | 363 | 139 |
| TOTA | 254673 | 203806 | 153179 | 167064 | 278640 | 238610 | 206006 | 198507 | 157295 |

Run title : Icelandic Haddock. Run 3.
At 24/04/2003 12:08
Terminal Fs derived using XSA (With F shrinkage)
Table 10 Stock number-at-age (start of year)
Numbers*10**-3

| YEAR | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | GMST82- <br> 00 | **AMST8 <br> $2-00$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |
| 2 | 98918 | 14495 | 48557 | 118479 | 155171 | 183881 | 0 | 50058 | 63338 |
| 3 | 27368 | 79743 | 11680 | 38780 | 94875 | 125042 | 149626 | 35429 | 46419 |
| 4 | 37513 | 19068 | 57951 | 8247 | 25873 | 67455 | 92781 | 25617 | 33687 |
| 5 | 11393 | 20645 | 10197 | 32157 | 4639 | 14737 | 40576 | 16488 | 21401 |
| 6 | 6407 | 4907 | 9311 | 3966 | 13826 | 2443 | 7425 | 8109 | 11724 |
| 7 | 9463 | 2948 | 1828 | 3116 | 1390 | 5699 | 982 | 3263 | 4911 |
| 8 | 1218 | 3241 | 1054 | 644 | 932 | 507 | 1837 | 1210 | 1977 |
| 9 | 95 | 371 | 949 | 331 | 198 | 325 | 193 | 412 | 748 |
| TOTAL | 192376 | 145419 | 141527 | 205720 | 296905 | 400088 | 293421 |  |  |

Table 3.4.6.2 (Cont'd)

Run title : Icelandic Haddock. Run 3.
At 24/04/2003 12:08
Table 16 Summary (without SOP correction)
Terminal Fs derived using XSA (With F shrinkage)

|  | RECRUITS | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | FBAR4-7 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Age2 |  |  |  |  |  |
| 1982 | 42212 | 198062 | 111696 | 69325 | 0.6207 | 0.4509 |
| 1983 | 30160 | 161861 | 101889 | 65943 | 0.6472 | 0.4738 |
| 1984 | 19932 | 125029 | 79753 | 48285 | 0.6054 | 0.4991 |
| 1985 | 41756 | 115975 | 59913 | 50933 | 0.8501 | 0.5154 |
| 1986 | 89202 | 114856 | 56359 | 48863 | 0.8670 | 0.7880 |
| 1987 | 167933 | 131172 | 41613 | 40801 | 0.9805 | 0.6395 |
| 1988 | 47637 | 161578 | 65960 | 54236 | 0.8223 | 0.6568 |
| 1989 | 26666 | 175049 | 99582 | 62979 | 0.6324 | 0.6592 |
| 1990 | 22362 | 151032 | 110530 | 67200 | 0.6080 | 0.5776 |
| 1991 | 80244 | 135817 | 91392 | 54732 | 0.5989 | 0.5964 |
| 1992 | 170353 | 133821 | 63383 | 47212 | 0.7449 | 0.6911 |
| 1993 | 37502 | 137297 | 69456 | 48844 | 0.7032 | 0.6804 |
| 1994 | 41275 | 135715 | 83215 | 59345 | 0.7132 | 0.6718 |
| 1995 | 70693 | 131819 | 86967 | 61131 | 0.7029 | 0.6545 |
| 1996 | 35049 | 113629 | 68509 | 56958 | 0.8314 | 0.7108 |
| 1997 | 98918 | 103023 | 61651 | 44053 | 0.7146 | 0.6219 |
| 1998 | 14495 | 97529 | 63533 | 41434 | 0.6522 | 0.6595 |
| 1999 | 48557 | 97845 | 62532 | 45481 | 0.7273 | 0.7178 |
| 2000 | 118479 | 105562 | 58152 | 42167 | 0.7251 | 0.7186 |
| 2001 | 155171 | 137018 | 59530 | 39647 | 0.6660 | 0.5749 |
| 2002 | 183881 | 191856 | 90775 | 50496 | 0.5563 | 0.6094 |
| Arith. |  |  |  |  | 0.6270 |  |
| Mean | 73451 | 135978 | 75542 | 52384 | 0.7128 |  |
| Units | (Thousands) | (Tonnes) | (Tonnes) | (Tonnes) |  |  |

Table 3.4.7.1 Haddock in division Va. Input file for RCT3.
Iceland Haddock: VPA and groundfish survey data

| 320 2 |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: |
| 'Yearcl' | VPAage2' | 'Surv3' | 'Surv2' | 'Surv1' |
| 1983 | 42 | 591 | 327 | -11 |
| 1984 | 89 | 1636 | 1085 | 282 |
| 1985 | 166 | 1848 | 2963 | 1240 |
| 1986 | 47 | 416 | 407 | 222 |
| 1987 | 26 | 273 | 234 | 158 |
| 1988 | 22 | 416 | 319 | 106 |
| 1989 | 79 | 1387 | 1460 | 705 |
| 1990 | 169 | 2451 | 2114 | 897 |
| 1991 | 37 | 398 | 377 | 181 |
| 1992 | 41 | 481 | 613 | 300 |
| 1993 | 70 | 1210 | 825 | 585 |
| 1994 | 35 | 509 | 663 | 359 |
| 1995 | 98 | 1083 | 1191 | 953 |
| 1996 | 14 | 258 | 182 | 86 |
| 1997 | 48 | 450 | 862 | 231 |
| 1998 | 117 | 1150 | 901 | 807 |
| 1999 | -11 | 2010 | 1481 | 606 |
| 2000 | -11 | 2828 | 2983 | 813 |
| 2001 | -11 | -11 | 979 | 211 |
| 2002 | -11 | -11 | -11 | 1120 |

Table 3.4.7.2 Haddock in division Va. Output from the RTC3 model. Analysis by RCT3 ver3.1 of data from file :

Recrun03.dat

Iceland Haddock: VPA and groundfish survey data
Data for 3 surveys over 20 years : 1983-2002
Regression type $=\mathrm{C}$
Tapered time weighting applied
power = 3 over 20 years
Survey weighting not applied
Final estimates shrunk towards mean
Minimum S.E. for any survey taken as . 20
Minimum of 3 points used for regression
Forecast/Hindcast variance correction used.
Year class $=1995$

| Survey/ <br> Series | Slope | Intercept | Std Error | Rsquare | $\begin{aligned} & \text { No. } \\ & \text { Pts } \end{aligned}$ | Index Value | Predicted Value | $\begin{aligned} & \text { Std } \\ & \text { Error } \end{aligned}$ | WAP <br> Weights |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Surv3 | . 94 | -2.20 | . 25 | . 887 | 12 | 6.99 | 4.37 | . 292 | . 406 |
| Surv2 | . 89 | -1.79 | . 27 | . 869 | 12 | 7.08 | 4.50 | . 320 | . 338 |
| Surv1 | . 98 | -1.68 | . 35 | . 807 | 11 | 6.86 | 5.02 | . 444 | . 175 |
|  |  |  |  |  | VPA | Mean = | 4.02 | . 657 | . 080 |

Year class = 1996

| Survey/ <br> Series | Slope | Intercept | $\begin{aligned} & \text { Std } \\ & \text { Error } \end{aligned}$ | Rsquare | $\begin{aligned} & \text { No. } \\ & \text { Pts } \end{aligned}$ | Index Value | Predicted Value | $\begin{gathered} \text { Std } \\ \text { Error } \end{gathered}$ | WAP <br> Weights |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Surv3 | . 97 | -2.35 | . 25 | . 881 | 13 | 5.56 | 3.02 | . 318 | . 386 |
| Surv2 | . 91 | -1.91 | . 26 | . 872 | 13 | 5.21 | 2.81 | . 344 | . 330 |
| Surv1 | . 93 | -1.45 | . 33 | . 812 | 12 | 4.47 | 2.70 | . 451 | . 192 |
|  |  |  |  |  | VPA | Mean = | 4.07 | . 647 | . 093 |

Year class = 1997

| Survey/ <br> Series | Slope | Intercept | $\begin{aligned} & \text { Std } \\ & \text { Error } \end{aligned}$ | Rsquare | $\begin{aligned} & \text { No. } \\ & \text { Pts } \end{aligned}$ | Index <br> Value | Predicted Value | $\begin{aligned} & \text { Std } \\ & \text { Error } \end{aligned}$ | WAP <br> Weights |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Surv3 | 1.04 | -2.84 | . 27 | . 893 | 14 | 6.11 | 3.50 | . 312 | . 328 |
| Surv2 | . 93 | -2.09 | . 25 | . 902 | 14 | 6.76 | 4.21 | . 294 | . 371 |
| Surv1 | . 92 | -1.43 | . 31 | . 864 | 13 | 5.45 | 3.61 | . 364 | . 242 |
|  |  |  |  |  | VPA | Mean = | 3.95 | . 734 | . 059 |
| Year class $=1998$ |  |  |  |  |  |  |  |  |  |
| I-----------Regression----------I I-----------Prediction---------- I |  |  |  |  |  |  |  |  |  |
| Survey/ | Slope | Inter- | Std | Rsquare | No. | Index | Predicted | Std | WAP |
| Series |  | cept | Error |  | Pts | Value | Value | Error | Weights |
| Surv3 | 1.04 | -2.82 | . 28 | . 873 | 15 | 7.05 | 4.50 | . 330 | . 309 |
| Surv2 | . 94 | -2.19 | . 26 | . 890 | 15 | 6.80 | 4.22 | . 299 | . 374 |
| Surv1 | . 92 | -1.37 | . 30 | . 858 | 14 | 6.69 | 4.77 | . 367 | . 249 |
|  |  |  |  |  | VPA | Mean = | 3.93 | . 703 | . 068 |

## Table 3.4.7.2 (Cont'd)

Year class $=1999$

| Survey/ <br> Series | Slope | Intercept | Std Error | Rsquare | $\begin{aligned} & \text { No. } \\ & \text { Pts } \end{aligned}$ | Index <br> Value | Predicted Value | $\begin{gathered} \text { Std } \\ \text { Error } \end{gathered}$ | WAP Weights |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Surv3 | 1.08 | -3.09 | . 29 | . 870 | 16 | 7.61 | 5.15 | . 362 | . 298 |
| Surv2 | 1.03 | -2.73 | . 32 | . 848 | 16 | 7.30 | 4.80 | . 378 | . 273 |
| Surv1 | . 92 | -1.37 | . 29 | . 875 | 15 | 6.41 | 4.50 | . 333 | . 352 |
|  |  |  |  |  | VPA | Mean $=$ | 3.99 | . 715 | . 076 |

Year class $=2000$

| Survey/ <br> Series | Slope | Intercept | $\begin{gathered} \text { Std } \\ \text { Error } \end{gathered}$ | Rsquare | $\begin{aligned} & \text { No. } \\ & \text { Pts } \end{aligned}$ | Index Value | Predicted Value | $\begin{aligned} & \text { Std } \\ & \text { Error } \end{aligned}$ | WAP <br> Weights |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Surv3 | 1.09 | -3.16 | . 29 | . 870 | 16 | 7.95 | 5.54 | . 395 | . 288 |
| Surv2 | 1.05 | -2.86 | . 32 | . 846 | 16 | 8.00 | 5.54 | . 435 | . 238 |
| Surv1 | . 91 | -1.36 | . 28 | . 879 | 15 | 6.70 | 4.75 | . 341 | . 387 |
|  |  |  |  |  | VPA | Mean = | 3.99 | . 718 | . 087 |

Year class = 2001


| Year <br> Class | Weighted <br> Average <br> Prediction | Log <br> WAP | Int <br> Std <br> Error | Ext <br> Std <br> Error | Var <br> Ratio | VPA | Log <br> VPA |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | ---: |
| 1995 | 89 | 4.50 | .19 | .15 | .69 | 99 | 4.60 |
| 1996 | 19 | 2.99 | .20 | .21 | 1.16 | 15 | 2.71 |
| 1997 | 45 | 3.81 | .18 | .19 | 1.07 | 48 | 3.89 |
| 1998 | 83 | 4.42 | .18 | .14 | .62 | 118 | 4.77 |
| 1999 | 113 | 4.74 | .20 | .20 | .98 |  |  |
| 2000 | 164 | 5.10 | .21 | .29 | 1.90 |  |  |
| 2001 | 48 | 3.89 | .24 | .28 | 1.42 |  |  |
| 2002 | 123 | 4.81 | .32 | .41 | 1.64 |  |  |

Table 3.4.8.1 Haddock Va. Input data for short-term prediction.
MFDP version 1
Run: had-iceg
Time and date: 17:28 07/05/03
Fbar age range: 4-7

| 2003 |  |  | Mat | PF | PM | SWt |  | Sel CWt |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N | M |  |  |  |  |  |  |  |
|  | 2 | 40129 | 0.2 | 0.062 | 0 | 0 | 0.23 | 1.63E-02 | 0.542 |
|  | 3 | 126781 | 0.2 | 0.347 | 0 | 0 | 0.412 | 0.135159 | 0.885 |
|  | 4 | 83629 | 0.2 | 0.685 | 0 | 0 | 0.801 | 0.352525 | 1.321 |
|  | 5 | 40738 | 0.2 | 0.867 | 0 | 0 | 1.268 | 0.555264 | 1.673 |
|  | 6 | 7355 | 0.2 | 0.922 | 0 | 0 | 1.873 | 0.746709 | 2.242 |
|  | 7 | 868 | 0.2 | 0.946 | 0 | 0 | 3.139 | 0.866501 | 2.656 |
|  | 8 | 1551 | 0.2 | 1 | 0 | 0 | 2.343 | 0.944449 | 2.814 |
|  | 9 | 134 | 0.2 | 1 | 0 | 0 | 3.301 | 0.944449 | 3.318 |
| 2004 |  |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM |  | SWt | Sel | CWt |
|  | 2 | 113545 | 0.2 | 0.07 | 0 | 0 | 0.197 | 1.63E-02 | 0.542 |
|  | 3 |  | 0.2 | 0.31 | 0 | 0 | 0.592 | 0.135159 | 0.873 |
|  | 4 |  | 0.2 | 0.61 | 0 | 0 | 0.758 | 0.352525 | 1.28 |
|  | 5 |  | 0.2 | 0.81 | 0 | 0 | 1.222 | 0.555264 | 1.737 |
|  | 6 |  | 0.2 | 0.89 | 0 | 0 | 1.754 | 0.746709 | 2.17 |
|  | 7 |  | 0.2 | 0.92 | 0 | 0 | 2.517 | 0.866501 | 2.702 |
|  | 8 |  | 0.2 | 0.97 | 0 | 0 | 2.646 | 0.944449 | 3.098 |
|  | 9 |  | 0.2 | 1 | 0 | 0 | 3.123 | 0.944449 | 3.318 |
| 2005 |  |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM |  | SWt | Sel | CWt |
|  | 2 | 51754 | 0.2 | 0.07 | 0 | 0 | 0.197 | 1.63E-02 | 0.542 |
|  | 3 |  | 0.2 | 0.31 | 0 | 0 | 0.502 | 0.135159 | 0.873 |
|  | 4. |  | 0.2 | 0.61 | 0 | 0 | 1.098 | 0.352525 | 1.265 |
|  | 5 |  | 0.2 | 0.81 | 0 | 0 | 1.175 | 0.555264 | 1.701 |
|  | 6 |  | 0.2 | 0.89 | 0 | 0 | 1.716 | 0.746709 | 2.237 |
|  | 7. |  | 0.2 | 0.92 | 0 | 0 | 2.375 | 0.866501 | 2.636 |
|  | 8. |  | 0.2 | 0.97 | 0 | 0 | 2.646 | 0.944449 | 3.129 |
|  | 9. |  | 0.2 | 1 | 0 | 0 | 3.123 | 0.944449 | 3.318 |

Input units are thousands and kg - output in tonnes

Table 3.4.8.2 Haddock in division Va. Input to yield-per-recruit.
MFYPR version 1
Run: had-iceg
Icelandic Haddock.
Time and date: 17:39 07/05/02
Fbar age range: 4-7

| Age | M | Mat | PF | PM | SWt | Sel | CWt |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | .2 | .068 | 0 | 0 | .1852 | .0162 | .54175 |
| 3 | .2 | .2605 | 0 | 0 | .4794 | .1265 | .8853 |
| 4 | .2 | .491 | 0 | 0 | .90845 | .36525 | 1.3237 |
| 5 | .2 | .6725 | 0 | 0 | 1.4003 | .57945 | 1.8088 |
| 6 | .2 | .79 | 0 | 0 | 1.96775 | .78455 | 2.3461 |
| 7 | .2 | .8635 | 0 | 0 | 2.51275 | .83915 | 2.97555 |
| 8 | .2 | .921 | 0 | 0 | 3.1646 | .97075 | 3.43515 |
| 9 | .2 | .9795 | 0 | 0 | 3.2482 | .86475 | 3.90525 |

Weights in kilograms

Table 3.4.8.3 Haddock in division Va. Output from yield-per-recruit.
MFYPR version 2 a
Run: had-iceg
Time and date: 19:55 07/05/02
Yield per results

| FMult | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNos <br> Jan | SSBJan | SpwnNos <br> Spwn |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0.0000 | 0.0000 | 0.0000 | 0.0000 | 5.5167 | 9.1405 | 3.2229 | 7.6094 | 3.2229 |  |
| 0.1000 | 0.0752 | 0.2003 | 0.5388 | 4.5198 | 6.1126 | 2.2847 | 4.7141 | 2.2847 |  |
| 0.2000 | 0.1504 | 0.3022 | 0.7389 | 4.0148 | 4.6597 | 1.8245 | 3.3531 | 1.8245 | 4.6094 |
| 0.3000 | 0.2255 | 0.3653 | 0.8234 | 3.7030 | 3.8149 | 1.5499 | 2.5795 | 1.5499 | 2.3531 |
| 0.4000 | 0.3007 | 0.4091 | 0.8598 | 3.4872 | 3.2653 | 1.3663 | 2.0884 | 1.3663 | 2.0884 |
| 0.5000 | 0.3759 | 0.4419 | 0.8741 | 3.3265 | 2.8802 | 1.2340 | 1.7524 | 1.2340 | 1.7524 |
| 0.6000 | 0.4511 | 0.4677 | 0.8776 | 3.2005 | 2.5953 | 1.1335 | 1.5099 | 1.1335 | 1.5099 |
| 0.7000 | 0.5262 | 0.4887 | 0.8758 | 3.0979 | 2.3760 | 1.0539 | 1.3275 | 1.0539 | 1.3275 |
| 0.8000 | 0.6014 | 0.5064 | 0.8712 | 3.0121 | 2.2017 | 0.9891 | 1.1857 | 0.9891 | 1.1857 |
| 0.9000 | 0.6766 | 0.5215 | 0.8653 | 2.9387 | 2.0597 | 0.9351 | 1.0726 | 0.9351 |  |
| 1.0000 | 0.7518 | 0.5347 | 0.8589 | 2.8749 | 1.9414 | 0.8891 | 0.9804 | 0.8891 | 1.0726 |
| 1.1000 | 0.8269 | 0.5464 | 0.8524 | 2.8185 | 1.8412 | 0.8494 | 0.9038 | 0.8494 | 0.9038 |
| 1.2000 | 0.9021 | 0.5569 | 0.8459 | 2.7683 | 1.7551 | 0.8146 | 0.8391 | 0.8146 | 0.8391 |
| 1.3000 | 0.9773 | 0.5663 | 0.8397 | 2.7230 | 1.6802 | 0.7839 | 0.7839 | 0.7839 | 0.7839 |
| 1.4000 | 1.0525 | 0.5749 | 0.8336 | 2.6819 | 1.6144 | 0.7565 | 0.7361 | 0.7565 | 0.7361 |
| 1.5000 | 1.1276 | 0.5827 | 0.8279 | 2.6443 | 1.5559 | 0.7318 | 0.6943 | 0.7318 | 0.6943 |
| 1.6000 | 1.2028 | 0.5900 | 0.8224 | 2.6097 | 1.5036 | 0.7094 | 0.6575 | 0.7094 | 0.6575 |
| 1.7000 | 1.2780 | 0.5967 | 0.8172 | 2.5777 | 1.4565 | 0.6890 | 0.6248 | 0.6890 | 0.6248 |
| 1.8000 | 1.3532 | 0.6029 | 0.8123 | 2.5480 | 1.4138 | 0.6703 | 0.5956 | 0.6703 | 0.5956 |
| 1.9000 | 1.4283 | 0.6088 | 0.8075 | 2.5202 | 1.3748 | 0.6532 | 0.5692 | 0.6532 | 0.5692 |
| 2.0000 | 1.5035 | 0.6142 | 0.8031 | 2.4943 | 1.3391 | 0.6372 | 0.5454 | 0.6372 | 0.5454 |


| Reference point | F multiplier | Absolute F |
| :--- | :--- | :--- |
| Fbar(4-7) | 1.0000 | 0.7518 |
| $\mathbf{F}_{\max }$ | 0.6045 | 0.4544 |
| $\mathbf{F}_{0.1}$ | 0.2384 | 0.1792 |
| F35\%SPR | 0.2865 | 0.2154 |

Weights in kilograms

Table 3.4.8.4 a Haddock in division Va. Output from short term prediction using ADCAM. Tac constraint of 65000 tonnes for 2003.

Output from short-term prognosis not exactly the same weights in 2005 as in MFDP.

| 2003 | Biom 3+ | Landings | $\mathrm{F}_{4-7}$ |
| :--- | :--- | :--- | :--- |
| SSB | 191 | 65 | 0.477 |
| 129 |  |  |  |


|  |  |  |  |  |  |  |  | 2005 |  |  |  |  |  | Biom 3+ | Landings | $\mathrm{F}_{4-7}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FMult | SSB | Biom 3+ | Landings | $\mathrm{F}_{4-7}$ | SSB | B |  |  |  |  |  |  |  |  |  |  |
| 0.15 | 148 | 204 | 19 | 0.095 | 208 | 281 | 26 | 0.095 |  |  |  |  |  |  |  |  |
| 0.2 | 148 | 204 | 25 | 0.126 | 203 | 275 | 34 | 0.126 |  |  |  |  |  |  |  |  |
| 0.25 | 148 | 204 | 31 | 0.158 | 198 | 270 | 41 | 0.158 |  |  |  |  |  |  |  |  |
| 0.3 | 148 | 204 | 36 | 0.189 | 194 | 265 | 47 | 0.189 |  |  |  |  |  |  |  |  |
| 0.35 | 148 | 204 | 42 | 0.221 | 190 | 260 | 53 | 0.221 |  |  |  |  |  |  |  |  |
| 0.4 | 148 | 204 | 47 | 0.252 | 185 | 255 | 58 | 0.252 |  |  |  |  |  |  |  |  |
| 0.45 | 148 | 204 | 52 | 0.284 | 181 | 250 | 63 | 0.284 |  |  |  |  |  |  |  |  |
| 0.5 | 148 | 204 | 57 | 0.315 | 177 | 246 | 67 | 0.315 |  |  |  |  |  |  |  |  |
| 0.55 | 148 | 204 | 62 | 0.347 | 174 | 241 | 71 | 0.347 |  |  |  |  |  |  |  |  |
| 0.6 | 148 | 204 | 67 | 0.378 | 170 | 237 | 75 | 0.378 |  |  |  |  |  |  |  |  |
| 0.65 | 148 | 204 | 72 | 0.41 | 166 | 232 | 78 | 0.41 |  |  |  |  |  |  |  |  |
| 0.7 | 148 | 204 | 77 | 0.441 | 163 | 228 | 81 | 0.441 |  |  |  |  |  |  |  |  |
| 0.75 | 148 | 204 | 81 | 0.473 | 159 | 224 | 84 | 0.473 |  |  |  |  |  |  |  |  |
| 0.8 | 148 | 204 | 85 | 0.504 | 156 | 220 | 87 | 0.504 |  |  |  |  |  |  |  |  |
| 0.85 | 148 | 204 | 90 | 0.536 | 152 | 217 | 89 | 0.536 |  |  |  |  |  |  |  |  |
| 0.9 | 148 | 204 | 94 | 0.567 | 149 | 213 | 91 | 0.567 |  |  |  |  |  |  |  |  |
| 0.95 | 148 | 204 | 98 | 0.599 | 146 | 209 | 93 | 0.599 |  |  |  |  |  |  |  |  |
| 1 | 148 | 204 | 102 | 0.63 | 143 | 206 | 94 | 0.63 |  |  |  |  |  |  |  |  |
| 1.05 | 148 | 204 | 106 | 0.662 | 140 | 202 | 96 | 0.662 |  |  |  |  |  |  |  |  |
| 1.1 | 148 | 204 | 109 | 0.693 | 137 | 199 | 97 | 0.693 |  |  |  |  |  |  |  |  |
| 1.15 | 148 | 204 | 113 | 0.725 | 135 | 196 | 98 | 0.725 |  |  |  |  |  |  |  |  |
| 1.2 | 148 | 204 | 117 | 0.756 | 132 | 193 | 99 | 0.756 |  |  |  |  |  |  |  |  |
| 1.25 | 148 | 204 | 120 | 0.788 | 129 | 189 | 100 | 0.788 |  |  |  |  |  |  |  |  |
| 1.3 | 148 | 204 | 124 | 0.819 | 127 | 186 | 101 | 0.819 |  |  |  |  |  |  |  |  |
| 1.35 | 148 | 204 | 127 | 0.851 | 124 | 184 | 101 | 0.851 |  |  |  |  |  |  |  |  |
| 1.4 | 148 | 204 | 130 | 0.882 | 122 | 181 | 102 | 0.882 |  |  |  |  |  |  |  |  |
| 1.45 | 148 | 204 | 133 | 0.914 | 119 | 178 | 102 | 0.914 |  |  |  |  |  |  |  |  |
| 1.5 | 148 | 204 | 136 | 0.945 | 117 | 175 | 103 | 0.945 |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 3.4.8.4 b Haddock in division Va. Output from short term prediction using MFDP. Tac constraint of 65000 tonnes for 2003.

MFDP version 1
Run: had-iceg
Index file 6/5/2002
Time and date: 17:28 07/05/03
Fbar age range: 4-7

| 2003 |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: |
| Biomass | SSB | FMult | FBar | Landings |
| 200683 | 128724 | 0.7593 | 0.4785 | 65000 |


| 2004 |  |  | 2005 |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar |  | Landings |  |  | Biomass |  | SSB |
| 226658 | 147805 | 0 | 0 | 0 | 302365 | 220461 |  |  |  |  |
| . | 147805 | 0.1 | 0.063 | 12733 | 290884 | 210480 |  |  |  |  |
| . | 147805 | 0.2 | 0.1261 | 24806 | 280025 | 201054 |  |  |  |  |
| . | 147805 | 0.3 | 0.1891 | 36260 | 269750 | 192150 |  |  |  |  |
| . | 147805 | 0.4 | 0.2521 | 47129 | 260023 | 183736 |  |  |  |  |
| . | 147805 | 0.5 | 0.3151 | 57447 | 250813 | 175782 |  |  |  |  |
| . | 147805 | 0.6 | 0.3782 | 67248 | 242087 | 168260 |  |  |  |  |
| . | 147805 | 0.7 | 0.4412 | 76559 | 233818 | 161145 |  |  |  |  |
| . | 147805 | 0.8 | 0.5042 | 85410 | 225979 | 154411 |  |  |  |  |
| . | 147805 | 0.9 | 0.5672 | 93825 | 218544 | 148037 |  |  |  |  |
| . | 147805 | 1 | 0.6303 | 101831 | 211490 | 142001 |  |  |  |  |
| . | 147805 | 1.1 | 0.6933 | 109450 | 204795 | 136282 |  |  |  |  |
| . | 147805 | 1.2 | 0.7563 | 116703 | 198438 | 130863 |  |  |  |  |
| . | 147805 | 1.3 | 0.8193 | 123610 | 192400 | 125726 |  |  |  |  |
| . | 147805 | 1.4 | 0.8824 | 130192 | 186663 | 120854 |  |  |  |  |
| . | 147805 | 1.5 | 0.9454 | 136465 | 181209 | 116233 |  |  |  |  |
| . | 147805 | 1.6 | 1.0084 | 142447 | 176022 | 111847 |  |  |  |  |
| . | 147805 | 1.7 | 1.0714 | 148153 | 171088 | 107683 |  |  |  |  |
| . | 147805 | 1.8 | 1.1345 | 153598 | 166393 | 103729 |  |  |  |  |
| . | 147805 | 1.9 | 1.1975 | 158797 | 161923 | 99972 |  |  |  |  |
| . | 147805 | 2 | 1.2605 | 163761 | 157665 | 96402 |  |  |  |  |

Input units are thousands and kg - output in tonnes


Figure 3.4.2.1 Haddock Division VA. Nominal landings (tonnes) 1905-2002.


Figure 3.4.3.1 Haddock in division Va. Age disaggregated catch in numbers plotted on log scale. The grey lines show $\mathrm{Z}=1$.


Figure 3.4.5.1 Haddock in division va. Total biomass index from the groundfish survey 1000 tonnes. The shaded area shows show the standard error in the estimate of the indices. Indices based on unweighed mean of all stations and number of stations with haddock times median of the haddock catch at those stations are shown for comparison.


Figure 3.4.5.2 Icelandic haddock. Total biomass indices from the groundfish surveys in March (lines and shading) and the groundfish survey in October vertical segments. The standard error in the estimate of the indices is shown in the figure.


Figure 3.4.5.3 Haddock in division Va. Survey indices plotted against survey indices of the same year class one year earlier. The letters in the figure are year classes. The dashed vertical lines show the most recent values.


Figure 3.4.5.4 Catchcurves from the groundfish survey. Grey lines show $\mathrm{Z}=1$.


Figure 3.4.5.5 Icelandic haddock. Survey indices vs. number in stock. Line fitted through origin on original scale. . The fitted line uses the data until 1999. Dashed lines show most recent estimates.


Figure 3.4.5.6 Percentage of survey index in the northern area


Figure 3.4.5.7 Catch per unit effort in the most important gear types. The figure is based on locations where more than $50 \%$ of the catch is haddock. A change occurred in the longline fleet starting September 1999. Earlier only vessels larger than 10 BRT were required to return logbooks but later all vessels were required to return logbooks.


Figure 3.4.5.8 Effort towards haddock. The effort is calculated as the ratio of the total landings for the gear and the CPUE based on records where haddock was more than $50 \%$ of the registered catch.

Landings


## Recruitment age 2



Fishing mortality hand harvest rate


Spawning stock biomass


Figure 3.4.6.1 Haddock in division Va. Summary plots.


Figure 3.4.6.2 Haddock in division Va. Model estimate of selection pattern and variance in survey and in the catch. Selection used in prognosis is the mean of last 5 years.



Figure 3.4.6.3 Haddock in division Va. Retrospective pattern from the ADCAM run using indices from age 1 to 9. The last 2 figures shows retrospective pattern with predictions 4 years ahead using the observed catches.


Figure 3.4.6.4 Residuals from the fit to survey data $\frac{\log \left(I_{a y}+\varepsilon_{a g e}\right)}{\log \left(I_{a y}+\varepsilon_{a g e}\right)}$ Coloured circles indicate positive residuals (observed $>$ modelled). The largest circle corresponds to a value of 0.78 and residuals are proportional to the area of the circles


Figure 3.4.6.5 Haddock in division Va. Observed (points) and modelled (lines) survey biomass.


Figure 3.4.6.6 Residuals from the fit to autumn survey data $\cdot \frac{\log \left(I_{a y}+\varepsilon_{a g e}\right)}{\log \left(I_{a y}+\varepsilon_{a g e}\right)}$ Coloured circles indicate positive residuals (observed > modelled). The largest circle corresponds to a value of 1 and residuals are proportional to the area of the circles


Figure 3.4.6.7 Haddock in division Va. Assessment using the autumn survey. Observed (points) and modelled (lines) survey biomass


Figure 3.4.8.1 Haddock in division Va. Yield-per-recruit.

Precautionary approach plot 1979-2002


Figure 3.4.8.2 Haddock in division Va. Spawning stock vs. fishing mortality. .


Figure 3.4.8.3 Haddock in division Va. Spawning stock vs. recruitment. . The labels in the figure show year classes.


Figure 3.4.8.3 (Cont'd) Haddock in division Va. Results from short-term simulations assuming fishing at $\mathrm{F}=0.47$ after 2003.


Figure 3.4.8.4 Haddock in division Va. Cumulative probability profiles of the catch in 2004 and 2005 assuming $\mathrm{F}=0.47$.

# THE COD STOCK COMPLEX IN GREENLAND (NAFO SUBAREA 1 AND ICES SUBAREA XIV) AND ICELANDIC WATERS (DIVISION VA) 

### 4.1 Inter-relationship Between the Cod Stocks in the Greenland-Iceland Area

Tagging experiments carried out at Greenland and Iceland show that mature cod at West Greenland migrate to East Greenland and Iceland (Tåning, 1937; Hansen, 1949; and Anon. 1971). The immature East Greenland cod seem not to emigrate to Iceland, but in some years immature cod migrate to the West Greenland stock (Anon. 1971). Tagging experiments at Iceland show that migration of mature cod from Iceland to Greenland waters occurs very seldom and can be ignored in stock assessments (Jonsson 1965, 1986). Migrations from Greenland waters to Iceland can, therefore, be regarded as a homing migration.

In egg and larval surveys cod eggs have been found in an almost continuos belt from Iceland to East Greenland, along the East Greenland coast, round Cape Farewell and over the banks at West Greenland (Tåning 1937, Anon. 1963). From 0 -group surveys carried out in the East Greenland-Iceland area since 1970, it becomes quite evident that the drift of 0group cod from the Iceland spawning grounds to the different nursery areas at Iceland varies from year to year. The same applies to the drift of 0 -group cod with the currents from Iceland to East Greenland (Table 4.1.1). In some years it seems that no larval drift has taken place to the Greenland area, while in other years some, and in some years like 1973 and 1984, considerable numbers drifted to East Greenland waters (Vílhjalmsson and Fridgeirsson 1976, Vílhjalmsson and Magnússon 1984, Sveinbjörnsson and Jónsson 1999). Since 1995, 0-group surveys were continued with the area coverage reduced to the Icelandic EEZ. However, the estimates of the year classes 1997 to 2002 are exceptionally high. In 2001, more than $60 \%$ of the 0 -group cod were distributed in northern areas off Iceland (Table 4.1.1) and an exceptional high proportion of 0 -group cod were distributed off East Iceland. However, none of these year classes seem to have drifted in significant numbers to the Greenland shelf.

The 1973 and 1984 year classes have been very important to the fisheries off both West and East Greenland. Tagging results have shown that when these two year classes became mature, they had migrated in large numbers from West to East Greenland and, to some extent, to the spawning area off the southwest coast of Iceland. This migration of mature cod from Greenland to Iceland influences the assessment of these stocks (Schopka, 1994) and it cannot therefore be ignored in the assessments.

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Schopka, S.A. 1994. Fluctuations in the cod stock off Iceland during the twentieth century in relation to changes in the fisheries and environment. In Jakobsson, J. et al. (Eds). Cod and Climate, ICES mar. Sci. Symp., 198: 175-193.

Tåning, Å.V. 1937. Some features in the migration of cod. J. Cons. int. Explor. Mer 12: 1-35.

Vilhjálmsson, H., and E. Friðgeirsson 1976. A review of 0-group surveys in the Iceland-East Greenland area in the years 1970-1975. ICES Coop. Res. Rep. 54: 1-34.

Vilhjálmsson, H., and J. V. Magnússon 1984. Report on the 0 -group fish survey in Icelandic and East-Greenland waters, August 1984. ICES C.M. 1984/H:66, 26 pp.

Table 4.1.1 Abundance indices of O-group cod from international and Icelandic O-group surveys (Sveinbjörnsson and Hjörleifsson, 2002) in the East Greenland/Iceland area, 1971-2002 (except 1972 and 1995-96).

| Year class | Dohrn Bank East Greenland | SE Iceland | SW Iceland | W Iceland | N Iceland | E Iceland | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | + | - | - | 60 | 214 | - | 283 |
| 1973 | 135 | 10 | 107 | 96 | 757 | 86 | 1191 |
| 1974 | 2 | - |  | 22 | 30 | + | 54 |
| 1975 | + | - | 2 | 50 | 73 | 5 | 130 |
| 1976 | 5 | 9 | 30 | 102 | 2015 | 584 | 2743 |
| 1977 | 7 | 2 | + | 26 | 305 | 94 | 435 |
| 1978 | 2 | - | + | 169 | 335 | 47 | 552 |
| 1979 | 2 | + | 1 | 22 | 345 | + | 370 |
| 1980 | 1 | 2 | + | 38 | 507 | 10 | 557 |
| 1981 | 19 | - | - | 41 | 19 | - | 78 |
| 1982 | + | - | + | 7 | 4 | - | 11 |
| 1983 | + | - | + | 85 | 66 | 2 | 153 |
| 1984 | 372 | 5 | + | 200 | 826 | 369 | 1772 |
| 1985 | 32 | $+$ | + | 581 | 197 | 2 | 812 |
| 1986 | + | 1 | 2 | 15 | 32 | + | 50 |
| 1987 | 7 | - | 1 | 2 | 61 | 10 | 81 |
| 1988 | 0 | - | 1 | 7 | 12 | + | 20 |
| 1989 | 1 | - | 3 | 7 | 30 | + | 41 |
| 1990 | 3 | - | + | 2 | 30 | 2 | 37 |
| 1991 | + | - | - | + | 5 | + | 6 |
| 1992 | 0 | - | + | 15 | 21 | 5 | 42 |
| 1993 | 1 | - | + | 36 | 116 | 2 | 155 |
| 1994 | 0 | - | 0 | 1 | 71 | 2 | 74 |
| 1997 | $4^{1}$ | + | + | 97 | 1007 | 46 | 1152 |
| $1998{ }^{2}$ |  | + | 2 | 814 | 1799 | 137 | 2752 |
| $1999{ }^{2}$ |  | 25 | 9 | 221 | 8255 | 898 | 9408 |
| $2000^{2}$ |  | 118 | 15 | 171 | 2520 | 264 | 3088 |
| $2001{ }^{2}$ |  | 55 | 0 | 38 | 1549 | 722 | 2364 |
| $2002{ }^{2}$ |  | 180 | 8 | 157 | 4106 | 702 | 5153 |

[^4]
### 5.1 Cod off Greenland (offshore component)

Prior to 1996, the cod stocks off Greenland have been divided into West and East Greenland or treated as one stock unit for assessment purposes to avoid migration effects. Fjord populations (inshore) have always been included. In 1996, the offshore component off West and East Greenland, the so called Bank Cod, was assessed separately as one stock unit and distinguished from the inshore populations for the first time. The completion of a re-evaluation of available German sampling data for the offshore catches back to 1955 enabled such an analysis given in the 1996 North-Western Working Group report (ICES 1996/Assess:15). Due to the severely depleted status of the offshore stock component, the directed cod fishery was given up in 1992, the final year in the VPA. Since then, no adequate data were available to update the assessment. Information on the historic VPA is available in ICES 2001/Assess:20. Therefore, the present report only includes updated survey results and catch information.

### 5.1.1 Trends in landings and fisheries

Officially reported catches are given in Tables 5.1.1 and 5.1.2 for West and East Greenland respectively and includes the inshore catches. Landings as used by the working group are listed in Table 5.1.3 by inshore areas for West Greenland and offshore areas for both West and East Greenland, their trends being illustrated in Fig. 5.1.1. In 1924 the offshore fishery at West Greenland took off and until 1929 the landings increased from 200 t to 22000 t and exceeded the level of 120000 t in 1931. The next 10 years landings were fluctuating in the range of $60000-130000 \mathrm{t}$ (Horsted 2000). During World War II catches decreased by $1 / 3$ as only Greenland and Portugal participated in the fishery. Less is known about cod fisheries at East Greenland waters, but since 1954 landing statistics have been available. In the next 15 years the East Greenland landings were only contributing between 2-10 \% of the total offshore landings. During a period from the mid 1950'ties to 1960 annual landings taken offshore averaged about 270000 t . In 1962 the offshore catches calumniated with landings of 440000 t . After this historic high landings decreased sharply by $90 \%$ to 46000 t in 1974 and even further down in 1977. The level of 40000 t was only exceeded during the periods 1982-83 and 19881990. Large changes in effort started in 1970, which increased during exploitation of the strong year classes born in 1973 and 1984. The offshore fishery was closed in 1986 and for the first 10 months in 1987. During 1989-90, the landings decreased from 85000 t by more than $85 \%$ to 11000 t . Since 1992 no directed cod fishery has taken place offshore in West Greenland, although very high quota are available. In the same period the reported landings varied between 120 t and 750 t in East Greenland. In 2002 a total offshore catch amounted to 448 t , less than $5 \%$ was reported as by-catch. No reports on discards have been available.

Miscellaneous gears, mainly long lines and gillnets, contributed $30-40 \%$ until 1977 but have disappeared since then ( ICES 2002/Assess:20). At the moment otter trawl board catches (OTB) are the only operating fishing gear and have been the most important throughout the time-series for offshore fisheries.

### 5.1.2 Surveys

### 5.1.2.1 Results of the German groundfish survey off West and East Greenland

Annual abundance and biomass indices have been derived using stratified random groundfish surveys covering shelf areas and the continental slope off West and East Greenland. Surveys commenced in 1982 and were primarily designed for the assessment of cod (Gadus morhua L.). A detailed description of the survey design and determination of these estimates was given in the report ICES 1993/Assess:18 and Working Doc. 3/2003. Figure 5.1.2 indicate names of the 14 strata, their geographic boundaries, depth ranges and areas in nautical square miles $\left(\mathrm{nm}^{2}\right)$. All strata were limited at the 3 mile line offshore except for some inshore regions off East Greenland where there is a lack of adequate bathymetric measurements. In 1984, 1992, and 1994 the survey coverage was incomplete off East Greenland and in 1995 and 2002 in West Greenland partly due to technical problems (Working Doc. 3/2003).

### 5.1.2.1.1 Stock abundance indices

Table 5.1.4 lists abundance and biomass indices for West and East Greenland, respectively and then combined for the years 1982-2002. Trends of the biomass estimates for West and East Greenland are shown in Figures 5.1.3, including the spawning stock. These Figure illustrate the pronounced increase in stock abundance and biomass indices from 23 million individuals and 45000 t in 1984 to 828 million individuals and 690000 t in 1987. This trend was the result of the recruitment of the predominating year classes 1984 and 1985, which were mainly distributed in the northern and the shallow strata off West Greenland during 1987-89. Such high indices were never observed in strata off East Greenland, although their abundance and biomass estimates increased during the period 1989-91 suggesting an eastward migration. During the period 1987-89, which were years with high abundance, the precision of survey indices was extremely low
due to enormous variation in catch per tow data. Since 1988, stock abundance and biomass indices decreased dramatically by $99 \%$ to only 5 million fish and 6000 t in 1993 . The 2002 survey results confirmed the severely depleted status of the stock, although they represent the highest stock size in 11 years (less than $5 \%$ of the abundance in 1987). The total abundance and biomass indices amounted to 14 million individuals and 22000 t , respectively, were $70 \%$ of the stock were distributed off East Greenland.

### 5.1.2.1.2 Age composition

Age disaggregated abundance indices for West, East Greenland and total are listed in Tables 5.1.5-7, respectively, and are based on 1242 individual age determinations. The recruiting year classes 1998-2001 are considered weak as compared to the strong 1984 and 1985 year classes. The year class 1999 at age 2 however is estimated as the third strongest year class since 1982 and thus to provide some recovery potential in the next few years. Indeed, at age 3 the 1999 year class was the most frequent age group in 2002. The 0 -and 1 -group indices are considered unrepresentative of year class strength at age 3 due to gear specifications while the age group 2 seems to be quantitatively estimated and to represent a reasonable recruitment index. (Figure 5.1.5).

### 5.1.2.1.3 Mean length-at-age

The trends of the mean length of the age groups 1-10 years for West and East Greenland are illustrated in Figure 5.1.6 and 5.1.7 respectively for the period 1982-2002. They reveal pronounced area and temperature effects. Age groups $2-$ 10 years off East Greenland were found to be significant longer than those off West Greenland. Driven by the high abundance of cod off West Greenland, weighted mean length and weight for the age groups $1-5$ displayed a decrease during 1986-87 and remained at low levels until 1991. Since then, the length-at-age at ages 3 to 8 years increased significantly and remained at that high level until 2000, when low values were recorded. The values for West Greenland illustrate a stable period the last three years. The 2002 values for East Greenland indicate a small decrease in length for the youngest age classes and a stable length for the older age classes. Mean weight-at-age can be obtained from regression $f(x)=0.00895 x^{3.00589}, X=$ length in cm , the equation has been determined on the basis of historic measurements.

### 5.1.2.2 Results of the Greenland groundfish survey off West Greenland

Since 1988, the Greenland Institute of Natural Resources has annually conducted a stratified-random trawl survey off West Greenland from July to September (Working Doc. 16/2003). The main purpose of the survey is to evaluate the biomass and abundance of Northern shrimp (Pandalus borealis), but since 1992 data on most fish spices have been recorded. The survey covers the offshore areas at West Greenland between $59^{\circ} 00^{\prime} \mathrm{N}$ and $72^{\circ} 30^{\prime} \mathrm{N}$ from the 3 -mile limit to the 600 m (Figure 5.1.8). The survey area is divided into 6 NAFO Divisions, and further subdivided into three depth strata ( $0-200,201-400$ and 401-600 m) on basis of depth contour lines. A minimum of two hauls per stratum is always planned. Due to lack of information of the bottom topography Div. 1AN and Disko Bay are considered as two single strata. The trawl is a Skjervoy 3000/20 with bobbin gear and double bag. The mesh size in the codend is 20 mm . and standard trawling time offshore is $15-30$ minutes at a mean towing speed of 2.5 knots. Cod smaller than approximately 20 cm are caught insufficiently due to the trawl distance between net and bottom. Stratified abundance and biomass estimates were calculated from catch-per-tow data using the stratum areas as weighting factor (Cochran, 1953). The coefficient of catchability was set at 1.0 , implying that estimates are merely indices of abundance and biomass. Confidence intervals (CI) were set at the $95 \%$ level of significance of the stratified mean.

### 5.1.2 2. $\quad$ Stock abundance indices

The biomass indices for cod were estimated to be between 4-7000t in the period 1988-1990. In 1992 the biomass decreased with more than $95 \%$ to only 217 t and remained at this low level until recent years. In 2001 a slight improvement was detected with a biomass index at a little more than 600 t and in 2002 the biomass level was estimated to be close to 2000 t figure 5.1.4. Abundance was estimated to be 4.3 millions which is the highest number in the timeseries (1992-2002) (Table 5.1.8 and 5.1.9).

### 5.1.2.2.2 Age composition

Age disaggregated abundance indices are listed in Table 5.1.10. In 2001, the recruiting year classes 1997, 1998 and 1999 dominated the stock by $94 \%$ with equal shares. In 2002 year class 1998 and 1999 contributed to nearly $80 \%$ of the total abundance. Their abundance at ages 3 and 4 represent highest values of the time-series. Age disaggregated abundance indices for West Greenland indicates occurrence of few year classes and a dominance of year class 1998 and 1999. In 2002 year class 2000 gave the secound highest value in the time serie. This indicates three year classes in row with high abundence. In 2002, age length keys were determined on the basis of 562 otoliths.

### 5.1.3 Biological sampling of commercial catches

No commercial sampling data were available to assess recent catch in numbers, weight and maturity-at-age.

### 5.1.4 State of the stock

A historic XSA tuning was run in 1996 with the final year as 1992 and the output is illustrated in Figures 5.1.8. and 5.1.9. The plots indicate the very high and fluctuating fishing mortality as well as periodic good year classes.

The two surveys, the German survey off West and East Greenland and the Greenland shrimp survey off West Greenland, do confirm that the offshore component of the cod off Greenland is at a very low level.

Both surveys indicate increased recruitment of the year classes 1997, 1998, 1999 and 2000, the year class 1999 being the third strongest at age 2 since 1982 in the German survey. However, the recruiting year classes cannot be described as strong being estimated to be less than $10 \%$ of the most recent previous strong year class of 1984. Although rebuilding to previous high stock sizes cannot be expected to occur based on these year classes, they suggest that the process of rebuilding may have begun.

The age composition of the stock indicates high mortality rates of juvenile cod for the past decade, especially off West Greenland.

### 5.1.5 Stock projection

Cod is described a common species in the Greenland fauna, although reaching here its ecological northern boundary. Given suitable environmental conditions, cod in the offshore areas of Greenland are considered to be self-sustaining. However, even with sizeable SSBs present and spawning occurring, water temperature may be so cold that eggs and larvae will not survive. Stock parameters, slow growth and poor conditions (Lloret and Rätz 2000), late maturation, and highly variable recruitment strongly affected by environmental conditions, suggest that to be sustainable, exploitation rates would need to be low, particularly in periods of cold water. In productive periods, higher exploitation rates could be sustainable, but it would be advisable to maintain a spawning stock biomass sufficiently large to buffer for brief periods of cold water.

The former VPA assessment of the offshore cod stocks off Greenland revealed that over-fishing was an important cause for the collapse of this unit in the beginning of the 70s. Since that time, the spawning stock has remained below 100000 $t$ and has not been able to produce adequate recruitment. Relatively strong year class were produced in 1973 and 1984 despite the low SSBs, but these are believed to have emigrated from Iceland as larvae. The migration back to Iceland as mature fish further diminished the contribution of those year classes to local egg production. Recruitment pulses from Iceland could contribute to a substantial recovery of the offshore component in the short-term. However, strong recruitment pulses are rare events ( 2 known occurrences in the last 30 years).

### 5.1.6 Estimation of management reference points

Total abundance from the German survey was rescaled to the historic VPA abundance by a linear equation on a log-log scale ( $\mathrm{r}^{2}=0.82$ ) Figure 5.1.11. The relationship between the abundance of $4+\operatorname{cod}$ and corresponding 3 years recruits indicates that abundance $4+$ should exceed more than 60 mill. before a decent recruitment is reached. In 2002 the rescaled numbers of $4+$ from the German survey was estimated to be 17.5 mill, which is considered beyond safe limits Figure 5.1.12.

### 5.1.7 Management considerations

No fishing should take place until a substantial increase in stock size is evident. Technical measures to avoid the bycatch of juvenile cod should be maintained (mandatory use of a 22 mm sorting grid since October 1, 2000).

### 5.1.8 Comments on the assessment

The present assessment is based on survey indices only, due to the termination of the cod directed offshore fishery in 1992.

The VPA assessment conducted in 1996 was affected by several uncertainties in data as well as ecological factors. The effect of emigration was only directly covered for the 1973 and 1984 year classes and had been taken into account by an increase of the natural mortality to 0.3 for age groups 5 and older. The sampling of commercial catches was historically rather inconsistent and did not cover the $30 \%$ taken by miscellaneous gears, mainly longlines and gillnets up to 1977 . Since 1991, catch-at-age and weight-at-age data had to be calculated using survey data. Maturity data were poorly reported implying uncertainties in spawning stock estimates.

No XSA tuning could be applied since 1997 when low levels in landings, effort and stock abundance were observed. The age disaggregated survey indices had to be adjusted to account for incomplete coverage of the survey area in 1992 and 1994.

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Table 5.1.1 Nominal catch (t) of Cod in NAFO Subarea 1, 1988-2002 as officially reported to ICES.

| Country | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | - | - | 51 | 1 | - | - |  |  |
| Germany | 6.574 | 12.892 | 7.515 | 96 | - | - | - |  |
| Greenland | 52.135 | 92.152 | 58.816 | 20.238 | 5.723 | 1.924 | 2.115 |  |
| Japan | 10 | - | - | - | - | - | - |  |
| Norway | 7 | 2 | 948 | - | - | - |  |  |
| UK | 927 | 3780 | 1.631 | - | - | - | - |  |
| Total | 59.653 | 108.826 | 68.961 | 20.335 | 5.723 | 1.924 | 2.115 |  |
| WG estimate | $62.653{ }^{2}$ | $111.567{ }^{3}$ | $98.474{ }^{4}$ | - | - | - |  |  |
| Country | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | $2002{ }^{1}$ |
| Faroe Islands | - | - | - |  |  |  |  |  |
| Germany | - | - | - |  |  |  |  |  |
| Greenland | 1.710 | 948 | 904 | 319 | 622 | 764 | 1680 | 3698 |
| Japan | - | - | - |  |  |  |  |  |
| Norway | - | - | - |  |  |  |  |  |
| UK | - | - | - |  |  |  |  |  |
| Total | 1.710 | 948 | 904 | 319 | 622 | 764 | 1680 | 3698 |
| WG estimate | - | - | - | - | - | - | - |  |

[^5]Table 5.1.2 Nominal catch ( t ) of cod in ICES Subarea XIV, 1988-2002 as officially reported to ICES.

| Country | 1988 | 1989 | 1990 | 1991 |  | 1992 | 1993 | 1994 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | 12 | 40 | - | - |  | - | - | 1 |
| Germany | 12.049 | 10.613 | 26.419 | 8.434 |  | 5.893 | 164 | 24 |
| Greenland | 345 | 3.715 | 4.442 | 6.677 |  | 1.283 | 241 | 73 |
| Iceland | 9 | - | - | - |  | 22 | - | - |
| Norway | - | - | 17 | 828 |  | 1.032 | 122 | 14 |
| Portugal |  |  |  |  |  |  |  |  |
| Russia |  | - | - | - |  | 126 |  | - |
| UK (Engl. and | - | 1.158 | 2.365 | 5.333 |  | 2.532 | - | - |
| Wales) |  |  |  |  |  |  |  |  |
| UK (Scotland) | - | 135 | 93 | 528 |  | 463 | 163 | - |
| United Kingdom | - | - | - | - |  | - | 46 | 296 |
| Total | 12.415 | 15.661 | 33.336 | 21.800 |  | 11.351 | - | 408 |
| WG estimate | $9.457{ }^{1}$ | $14.669{ }^{2}$ | $33.513^{3}$ | $21.818{ }^{4}$ |  | - | 736 | - |
| - |  |  |  |  |  |  |  |  |
| Country | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | $2002{ }^{5}$ |
| Faroe Islands | - | - | - | - | 6 |  |  | 164 |
| Germany | 22 | 5 | 39 | 128 | 13 | 3 | 92 | 5 |
| Greenland | 29 | 5 | 32 | $37{ }^{5}$ | $+{ }^{5}$ |  | 4 | 232 |
| Iceland | 1 | - | - |  | - | - | 210 |  |
| Norway | + | 1 | - | + | 2 | - ${ }^{5}$ | 43 | 13 |
| Portugal |  |  |  | 31 | - | - | 278 |  |
| Russia | - | - | - |  |  |  |  |  |
| UK (E/W/NI) | 232 | 181 | 284 | 149 | 95 | 149 | 129 |  |
| UK (Scotland) | - | - | - |  |  |  |  |  |
| United Kingdom |  |  |  |  |  |  |  | 34 |
| Total | 284 | 192 | 355 | 345 | 116 | 152 | 756 | 448 |
| WG estimate | - | - | - | - | - | - |  |  |
| ${ }^{1}$ ) Excluding 3,000 t assumed to be from NAFO Division 1F and including 42 t taken by Japan |  |  |  |  |  |  |  |  |
| ${ }^{2}$ ) Excluding 2,741 tassumed to be from NAFO Division 1F and including 1,500 t reported from other areas assumed to be from Subarea XIV and including 94 t by Japan and 155 t by Greenland (Horsted, 1994) |  |  |  |  |  |  |  |  |
| ${ }^{3}$ ) Includes 129 t by Japan and 48 t additional catches by Greenland (Horsted, 1994) |  |  |  |  |  |  |  |  |
| ${ }^{4}$ ) Includes 18 t by Japan |  |  |  |  |  |  |  |  |

Table 5.1.3 Cod off Greenland. Catches ( t ) from 1924 - 2002 as used by the Working Group, inshore and offshore by NAFO div 1Band 1D offshore divided into East and West Greenland. Based on Horsted (1994, 2000).

| Cod | Inshore |  | Offshore |  |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Nafo 1 B | Nafo 1D | Total inshore | East | West | Total offshore | Greenland |
| 1924 | 131 | 221 | 843 |  | 200 | 200 | 1043 |
| 1925 | 122 | 318 | 1024 |  | 1871 | 1871 | 2895 |
| 1926 | 97 | 673 | 2224 |  | 4452 | 4452 | 6676 |
| 1927 | 282 | 982 | 3570 |  | 4427 | 4427 | 7997 |
| 1928 | 426 | 1153 | 4163 |  | 5871 | 5871 | 10034 |
| 1929 | 1479 | 1335 | 7080 |  | 22304 | 22304 | 29384 |
| 1930 | 2208 | 1681 | 9658 |  | 94722 | 94722 | 104380 |
| 1931 | 1905 | 1520 | 9054 |  | 120858 | 120858 | 129912 |
| 1932 | 1713 | 1042 | 9232 |  | 87273 | 87273 | 96505 |
| 1933 | 1799 | 1148 | 8238 |  | 54351 | 54351 | 62589 |
| 1934 | 2080 | 952 | 9468 |  | 88122 | 88122 | 97590 |
| 1935 | 1870 | 769 | 7526 |  | 65846 | 65846 | 73372 |
| 1936 | 2039 | 705 | 7174 |  | 125972 | 125972 | 133146 |
| 1937 | 1982 | 854 | 6961 |  | 90296 | 90296 | 97257 |
| 1938 | 1743 | 703 | 5492 |  | 90042 | 90042 | 95534 |
| 1939 | 2256 | 896 | 7161 |  | 89807 | 89807 | 96968 |
| 1940 | 2478 | 1061 | 8026 |  | 43122 | 43122 | 51148 |
| 1941 | 3229 | 823 | 8622 |  | 35000 | 35000 | 43622 |
| 1942 | 3831 | 1332 | 12027 |  | 40814 | 40814 | 52841 |
| 1943 | 5056 | 1240 | 13026 |  | 47400 | 47400 | 60426 |
| 1944 | 4322 | 1547 | 13385 |  | 51627 | 51627 | 65012 |
| 1945 | 4987 | 1207 | 14289 |  | 45800 | 45800 | 60089 |
| 1946 | 5210 | 1438 | 15262 |  | 44395 | 44395 | 59657 |
| 1947 | 5261 | 2096 | 18029 |  | 63458 | 63458 | 81487 |
| 1948 | 5660 | 1657 | 18675 |  | 109058 | 109058 | 127733 |
| 1949 | 4580 | 2110 | 17050 |  | 156015 | 156015 | 173065 |
| 1950 | 6358 | 2357 | 21173 |  | 179398 | 179398 | 200571 |
| 1951 | 5322 | 2571 | 18200 |  | 222340 | 222340 | 240540 |
| 1952 | 4443 | 2437 | 16726 |  | 317545 | 317545 | 334271 |
| 1953 | 5030 | 5513 | 22651 |  | 225017 | 225017 | 247668 |
| 1954 | 6164 | 3275 | 18698 | 4321 | 286120 | 290441 | 309139 |
| 1955 | 5523 | 4061 | 19787 | 5135 | 247931 | 253066 | 272853 |
| 1956 | 5373 | 5127 | 21028 | 12887 | 302617 | 315504 | 336532 |
| 1957 | 6146 | 5257 | 24593 | 10453 | 246042 | 256495 | 281088 |
| 1958 | 6178 | 5456 | 25802 | 10915 | 294119 | 305034 | 330836 |
| 1959 | 6404 | 5009 | 27577 | 19178 | 207665 | 226843 | 254420 |
| 1960 | 6741 | 3614 | 27099 | 23914 | 215737 | 239651 | 266750 |
| 1961 | 6569 | 4178 | 33965 | 19690 | 313626 | 333316 | 367281 |
| 1962 | 7809 | 3824 | 35380 | 17315 | 425278 | 442593 | 477973 |
| 1963 | 4877 | 2804 | 23269 | 23057 | 405441 | 428498 | 451767 |
| 1964 | 3311 | 8766 | 21986 | 35577 | 327752 | 363329 | 385315 |
| 1965 | 5209 | 6046 | 24322 | 17497 | 342395 | 359892 | 384214 |
| 1966 | 8738 | 7022 | 29076 | 12870 | 339130 | 352000 | 381076 |
| 1967 | 5658 | 6747 | 27524 | 24732 | 401955 | 426687 | 454211 |
| 1968 | 1669 | 6123 | 20587 | 15701 | 373013 | 388714 | 409301 |
| 1969 | 1767 | 7540 | 21492 | 17771 | 193163 | 210934 | 232426 |
| 1970 | 1469 | 3661 | 15613 | 20907 | 97891 | 118798 | 134411 |
| 1971 | 1807 | 3802 | 13506 | 32616 | 107674 | 140290 | 153796 |
| 1972 | 1855 | 3973 | 14645 | 26629 | 95974 | 122603 | 137248 |
| 1973 | 1362 | 3682 | 9622 | 11752 | 53320 | 65072 | 74694 |
| 1974 | 926 | 2588 | 8638 | 6553 | 39396 | 45949 | 54587 |

Table 5.1.3
Cod off Greenland. Continued.

| Year | Nafo 1 B | Nafo 1D | Total inshore | East | West | Total offshore | Greenland |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 1038 | 1269 | 6557 | 5925 | 41352 | 47277 | 53834 |
| 1976 | 644 | 904 | 5174 | 13027 | 28114 | 41141 | 46315 |
| 1977 | 580 | 2946 | 13999 | 8775 | 23997 | 32772 | 46771 |
| 1978 | 1587 | 2614 | 19679 | 7827 | 18852 | 26679 | 46358 |
| 1979 | 1768 | 6378 | 35590 | 8974 | 12315 | 21289 | 56879 |
| 1980 | 2303 | 7781 | 38571 | 11244 | 8291 | 19535 | 58106 |
| 1981 | 2810 | 6119 | 39703 | 10381 | 13753 | 24134 | 63837 |
| 1982 | 2448 | 7186 | 26664 | 20929 | 30342 | 51271 | 77935 |
| 1983 | 2803 | 7330 | 28652 | 13378 | 27825 | 41203 | 69855 |
| 1984 | 3908 | 5414 | 19958 | 8914 | 13458 | 22372 | 42330 |
| 1985 | 2936 | 1976 | 8441 | 2112 | 6437 | 8549 | 16990 |
| 1986 | 1038 | 1209 | 5302 | 4755 | 1301 | 6056 | 11358 |
| 1987 | 2995 | 8110 | 18486 | 6909 | 3937 | 10846 | 29332 |
| 1988 | 6294 | 2992 | 18791 | 12457 | 36824 | 49281 | 68072 |
| 1989 | 8491 | 8212 | 38529 | 15910 | 70295 | 86205 | 124734 |
| 1990 | 9857 | 9826 | 28799 | 33508 | 40162 | 73670 | 102469 |
| 1991 | 8641 | 2782 | 18311 | 21596 | 2024 | 23620 | 41931 |
| 1992 | 2710 | 1070 | 5723 | 11349 | 4 | 11353 | 17076 |
| 1993 | 323 | 968 | 1924 | 1135 | 0 | 1135 | 3059 |
| 1994 | 332 | 914 | 2115 | 437 | 0 | 437 | 2552 |
| 1995 | 521 | 332 | 1710 | 284 | 0 | 284 | 1994 |
| 1996 | 211 | 164 | 948 | 192 | 0 | 192 | 1140 |
| 1997 | 446 | 99 | 1186 | 370 | 0 | 370 | 1556 |
| 1998 | 118 | 78 | 323 | 346 | 0 | 346 | 669 |
| 1999 | 142 | 336 | 622 | 112 | 0 | 112 | 734 |
| 2000 | 266 | 332 | 764 | 100 | 0 | 100 | 864 |
| 2001 | 1183 | 54 | 1680 | 221 | 0 | 221 | 1901 |
| 2002 | 1803 | 214 | 3698* | 448 | 0 | 448 | 4146* |

Table 5.1.4 Cod off Greenland (offshore component), German survey. Abundance (1000) and biomass indices (t) for West, East Greenland and total by stratum, 1982-2002. Confidence intervals (CI) are given in per cent of the stratified mean at $95 \%$ level of significance. () incorrect due to incomplete sampling.

| YEAR | Abundance |  |  |  |  | Biomass |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | WEST | EAST | TOTAL | CI | Spawn. St. | WEST | EAST | TOTAL | CI | Spawn. St. |
| 1982 | 92276 | 8090 | 100366 | 28 | 33793 | 128491 | 23617 | 152107 | 25 | 79511 |
| 1983 | 50204 | 7991 | 58195 | 25 | 23889 | 82374 | 34157 | 116531 | 25 | 57223 |
| 1984 | 16684 | (6603) | (23286) | 32 | 17653 | 25566 | (19744) | (45309) | 34 | 36162 |
| 1985 | 59343 | 12404 | 71747 | 33 | 17349 | 35672 | 33565 | 69236 | 39 | 45630 |
| 1986 | 145682 | 15234 | 160915 | 32 | 14350 | 86719 | 41185 | 127902 | 26 | 48976 |
| 1987 | 786392 | 41635 | 828026 | 59 | 25467 | 638588 | 51592 | 690181 | 63 | 65584 |
| 1988 | 626493 | 23588 | 650080 | 48 | 128578 | 607988 | 52946 | 660935 | 46 | 155556 |
| 1989 | 358725 | 91732 | 450459 | 59 | 332589 | 333850 | 239546 | 573395 | 46 | 514773 |
| 1990 | 34525 | 25254 | 59777 | 43 | 46355 | 34431 | 65964 | 100395 | 34 | 77064 |
| 1991 | 4805 | 10407 | 15213 | 29 | 6404 | 5150 | 32751 | 37901 | 36 | 17756 |
| 1992 | 2043 | (658) | (2700) | 50 | 560 | 607 | (1216) | (1823) | 69 | 1091 |
| 1993 | 1437 | 3301 | 4738 | 36 | 2327 | 359 | 5600 | 5959 | 41 | 4024 |
| 1994 | 574 | (801) | (1375) | 36 | 457 | 140 | (2792) | (2930) | 68 | 1732 |
| 1995 | 278 | 7187 | 7463 | 93 | 2340 | 57 | 15525 | 15581 | 155 | 10445 |
| 1996 | 811 | 1447 | 2257 | 38 | 592 | 373 | 3599 | 3973 | 56 | 2017 |
| 1997 | 315 | 4153 | 4469 | 75 | 3411 | 284 | 13722 | 14007 | 90 | 10416 |
| 1998 | 1723 | 1671 | 3394 | 54 | 1133 | 130 | 4348 | 4479 | 91 | 3820 |
| 1999 | 912 | 2769 | 3681 | 34 | 809 | 240 | 3917 | 4157 | 62 | 3004 |
| 2000 | 1926 | 4816 | 6742 | 36 | 3556 | 570 | 4778 | 5349 | 40 | 4176 |
| 2001 | 8160 | 7604 | 15764 | 39 | 8252 | 2666 | 15271 | 17937 | 42 | 13381 |
| 2002 | 4121 | 9691 | 13812 | 41 | 11689 | 2110 | 19726 | 21836 | 51 | 21299 |

Table 5.1.5 Cod off West Greenland (offshore component), German survey. Age disaggregate abundance indices (1000), 1982-2002.*) calculated proportionally using age compositions reported by the ICES Working Group on Cod Stocks off East Greenland (ICES 1984/Assess:5).

| YEAR | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | $11+$ | TOTAL |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1982 | 0 | 176 | 884 | 33470 | 11368 | 32504 | 9528 | 2622 | 578 | 939 | 91 | 90 | 92250 |
| $* 1983$ | 0 | 0 | 1469 | 2815 | 26619 | 4960 | 10969 | 1882 | 992 | 317 | 168 | 13 | 50204 |
| 1984 | 159 | 5 | 38 | 2070 | 1531 | 9848 | 842 | 1873 | 87 | 186 | 27 | 0 | 16666 |
| 1985 | 831 | 38016 | 1481 | 948 | 6403 | 2833 | 7682 | 467 | 646 | 27 | 35 | 0 | 59369 |
| 1986 | 0 | 14148 | 112532 | 4089 | 903 | 6823 | 2095 | 4271 | 133 | 616 | 34 | 39 | 145683 |
| 1987 | 0 | 317 | 45473 | 692567 | 24230 | 5929 | 11813 | 1637 | 4006 | 0 | 366 | 30 | 786368 |
| 1988 | 0 | 257 | 3332 | 102767 | 510980 | 5425 | 613 | 1122 | 654 | 1274 | 32 | 35 | 626491 |
| 1989 | 12 | 204 | 2461 | 3565 | 93687 | 254002 | 3934 | 0 | 535 | 114 | 228 | 0 | 358742 |
| 1990 | 159 | 47 | 1007 | 3005 | 1244 | 21724 | 7221 | 47 | 0 | 0 | 0 | 19 | 34473 |
| 1991 | 0 | 293 | 224 | 476 | 1397 | 164 | 1894 | 317 | 6 | 0 | 0 | 0 | 4771 |
| 1992 | 0 | 263 | 1427 | 220 | 36 | 77 | 0 | 28 | 0 | 0 | 0 | 0 | 2051 |
| 1993 | 0 | 10 | 832 | 544 | 20 | 28 | 6 | 0 | 0 | 0 | 0 | 0 | 1440 |
| 1994 | 0 | 283 | 45 | 199 | 38 | 5 | 0 | 5 | 0 | 0 | 0 | 0 | 575 |
| 1995 | 0 | 0 | 241 | 16 | 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 279 |
| 1996 | 0 | 147 | 11 | 638 | 10 | 0 | 10 | 0 | 0 | 0 | 0 | 0 | 816 |
| 1997 | 0 | 12 | 27 | 15 | 263 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 317 |
| 1998 | 48 | 1642 | 0 | 0 | 5 | 25 | 0 | 0 | 0 | 0 | 0 | 0 | 1720 |
| 1999 | 29 | 401 | 392 | 87 | 7 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 922 |
| 2000 | 0 | 165 | 1015 | 615 | 116 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1911 |
| 2001 | 0 | 620 | 6202 | 1100 | 159 | 51 | 0 | 0 | 0 | 0 | 0 | 0 | 8132 |
| 2002 | 12 | 13 | 1061 | 2972 | 64 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4122 |

Table 5.1.6 Cod off East Greenland (offshore component), German survey. Age disaggregate abundance indices (1000), 1982-2002. *) calculated proportionally using age compositions reported by the ICES Working Group on Cod Stocks off East Greenland (ICES 1984/Assess:5). () incomplete sampling.

| YEAR | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | $11+$ | TOTAL |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1982 | 0 | 0 | 239 | 841 | 1764 | 1999 | 1227 | 379 | 130 | 1392 | 73 | 72 | 8116 |
| $* 1983$ | 0 | 0 | 411 | 605 | 1008 | 1187 | 2125 | 1287 | 302 | 265 | 703 | 101 | 7994 |
| $(1984)$ | 0 | 18 | 74 | 1342 | 657 | 1397 | 855 | 1617 | 407 | 103 | 36 | 95 | 6601 |
| 1985 | 230 | 1932 | 556 | 118 | 2494 | 2034 | 1852 | 785 | 2000 | 295 | 56 | 36 | 12388 |
| 1986 | 0 | 1397 | 3351 | 1693 | 551 | 2417 | 1120 | 2191 | 566 | 1627 | 116 | 139 | 15168 |
| 1987 | 0 | 13 | 13785 | 17788 | 3890 | 1027 | 1770 | 457 | 1571 | 187 | 1093 | 36 | 41617 |
| 1988 | 1 | 25 | 163 | 6982 | 11094 | 2016 | 480 | 1435 | 152 | 674 | 98 | 469 | 23599 |
| 1989 | 0 | 7 | 179 | 489 | 17396 | 63216 | 3021 | 294 | 4870 | 406 | 1795 | 42 | 91715 |
| 1990 | 0 | 38 | 80 | 551 | 462 | 5128 | 18012 | 265 | 72 | 251 | 0 | 349 | 25208 |
| 1991 | 0 | 106 | 377 | 394 | 685 | 147 | 3512 | 5035 | 81 | 37 | 11 | 9 | 10394 |
| $(1992)$ | 15 | 44 | 77 | 74 | 69 | 54 | 47 | 143 | 52 | 0 | 0 | 6 | 581 |
| 1993 | 0 | 17 | 44 | 1857 | 370 | 279 | 278 | 88 | 272 | 95 | 0 | 0 | 3300 |
| $(1994)$ | 0 | 87 | 0 | 29 | 261 | 143 | 87 | 145 | 0 | 29 | 0 | 0 | 781 |
| 1995 | 0 | 7 | 2523 | 1125 | 370 | 1730 | 450 | 141 | 460 | 36 | 217 | 125 | 7184 |
| 1996 | 0 | 0 | 0 | 502 | 258 | 295 | 255 | 60 | 77 | 0 | 0 | 0 | 1447 |
| 1997 | 0 | 0 | 37 | 28 | 1508 | 1611 | 566 | 236 | 140 | 0 | 0 | 19 | 4145 |
| 1998 | 63 | 240 | 192 | 21 | 45 | 462 | 435 | 156 | 43 | 0 | 0 | 0 | 1657 |
| 1999 | 191 | 632 | 665 | 417 | 138 | 302 | 179 | 200 | 0 | 35 | 24 | 0 | 2783 |
| 2000 | 0 | 808 | 1074 | 1341 | 787 | 157 | 291 | 75 | 141 | 115 | 31 | 0 | 4820 |
| 2001 | 0 | 309 | 944 | 1468 | 2244 | 1349 | 705 | 211 | 191 | 73 | 36 | 9 | 7539 |
| 2002 | 96 | 8 | 415 | 1824 | 2026 | 2080 | 1952 | 889 | 235 | 83 | 36 | 30 | 9674 |

Table 5.1.7 Cod off Greenland (total offshore component), German survey. Age disaggregate abundance indices (1000), 1982-2002. *) calculated proportionally using age compositions reported by the ICES Working Group on Cod Stocks off East Greenland (ICES 1984/Assess:5). () incomplete sampling.

| YEAR | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | $11+$ | TOTAL |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1982 | 0 | 176 | 1123 | 34311 | 13132 | 34503 | 10755 | 3001 | 708 | 2331 | 164 | 162 | 100366 |
| $* 1983$ | 0 | 0 | 1880 | 3420 | 27627 | 6147 | 13094 | 3169 | 1294 | 582 | 871 | 1140 | 58198 |
| $(1984)$ | 159 | 23 | 112 | 3412 | 2188 | 11245 | 1697 | 3490 | 494 | 289 | 63 | 95 | 23267 |
| 1985 | 1061 | 39948 | 2037 | 1066 | 8897 | 4867 | 9534 | 1252 | 2646 | 322 | 91 | 36 | 71757 |
| 1986 | 0 | 15545 | 115883 | 5782 | 1454 | 9240 | 3215 | 6462 | 699 | 2243 | 150 | 178 | 160851 |
| 1987 | 0 | 330 | 59258 | 710355 | 28120 | 6956 | 13583 | 2094 | 5577 | 187 | 1459 | 66 | 827985 |
| 1988 | 11 | 282 | 3495 | 109749 | 522074 | 7441 | 1093 | 2557 | 806 | 1948 | 130 | 504 | 650090 |
| 1989 | 12 | 211 | 2640 | 4054 | 111083 | 317218 | 6955 | 294 | 5405 | 520 | 2023 | 42 | 450457 |
| 1990 | 159 | 85 | 1087 | 3556 | 1706 | 26852 | 25233 | 312 | 72 | 251 | 0 | 368 | 59681 |
| 1991 | 0 | 399 | 601 | 870 | 2082 | 311 | 5406 | 5352 | 87 | 37 | 11 | 9 | 15165 |
| $(1992)$ | 15 | 307 | 1504 | 294 | 105 | 131 | 47 | 171 | 52 | 0 | 0 | 6 | 2632 |
| 1993 | 0 | 27 | 876 | 2401 | 390 | 307 | 284 | 88 | 272 | 95 | 0 | 0 | 4740 |
| $(1994)$ | 0 | 370 | 45 | 228 | 299 | 148 | 87 | 150 | 0 | 29 | 0 | 0 | 1356 |
| 1995 | 0 | 7 | 2764 | 1141 | 392 | 1730 | 450 | 141 | 460 | 36 | 217 | 125 | 7463 |
| 1996 | 0 | 147 | 11 | 1140 | 268 | 295 | 265 | 60 | 77 | 0 | 0 | 0 | 2263 |
| 1997 | 0 | 12 | 64 | 43 | 1771 | 1611 | 566 | 236 | 140 | 0 | 0 | 19 | 4462 |
| 1998 | 111 | 1882 | 192 | 21 | 50 | 487 | 435 | 156 | 43 | 0 | 0 | 0 | 3377 |
| 1999 | 220 | 1033 | 1057 | 504 | 145 | 302 | 185 | 200 | 0 | 35 | 24 | 0 | 3705 |
| 2000 | 0 | 973 | 2089 | 1956 | 903 | 157 | 291 | 75 | 141 | 115 | 31 | 0 | 6731 |
| 2001 | 0 | 929 | 7146 | 2568 | 2403 | 1400 | 705 | 211 | 191 | 73 | 36 | 9 | 15671 |
| 2002 | 108 | 21 | 1476 | 4796 | 2090 | 2080 | 1952 | 889 | 235 | 83 | 36 | 30 | 13796 |

Table 5.1.8 Cod off Greenland (offshore component), Greenland survey. Abundance indices (1000) for West Greenland by stratum, 1991-2002. Confidence intervals (CI) are given in percent of the stratified mean at $95 \%$ level of significance. () incorrect due to incomplete sampling.

| Year | 1 AN | 1 AS | 1AX | 1BN | 1BS | 1 C | 1 D | 1 E | 1 F | West. | CI |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1991 | $*$ | 0 | 11 | 7 | 32 | 429 | 78 | $*$ | $*$ | $(528)$ | 73 |
| 1992 | 0 | 0 | 4 | 16 | 33 | 242 | 242 | 0 | 9 | 547 | 45 |
| 1993 | 0 | 0 | 0 | 0 | 0 | 54 | 36 | 205 | 12 | 308 | 67 |
| 1994 | 9 | 0 | 0 | 0 | 54 | 98 | 0 | 7 | 0 | 167 | 43 |
| 1995 | 0 | 0 | 0 | 33 | 17 | 504 | 42 | 20 | 46 | 662 | 58 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 47 | 78 | 66 | 108 | 298 | 40 |
| 1997 | 0 | 0 | 0 | 2 | 8 | 35 | 0 | 0 | 0 | 45 | 64 |
| 1998 | 0 | 0 | 0 | 5 | 0 | 0 | 25 | 28 | 4 | 62 | 44 |
| 1999 | 0 | 10 | 18 | 141 | 52 | 17 | 18 | 8 | 0 | 261 | 41 |
| 2000 | 0 | 188 | 273 | 311 | 201 | 86 | 47 | 9 | 205 | 1321 | 19 |
| 2001 | 0 | 0 | 15 | 249 | 86 | 140 | 498 | 210 | 373 | 1570 | 23 |
| 2002 | 0 | 0 | 9 | 75 | 172 | 99 | 3595 | 102 | 202 | 4254 | 52 |

Table 5.1.9 Cod off Greenland (offshore component), Greenland survey. Biomass indices (t) for West Greenland by stratum, 1988-2002. Confidence intervals (CI) are given in per cent of the stratified mean at $95 \%$ level of significance. () incorrect due to incomplete sampling.

| Year | 1 AN | 1 AS | 1 AX | 1BN | 1 BS | 1 C | 1 D | 1 E | 1 F | West. | CI |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1988 | 0 | 0 | $*$ | 35 | 0 | 1230 | 2613 | $*$ | $*$ | $(3879)$ | 81 |
| 1989 | 44 | 0 | $*$ | 73 | 0 | 41 | 1002 | $*$ | $*$ | $(1217)$ | 51 |
| 1990 | 4 | 13 | $*$ | 7 | 7 | 118 | 6825 | $*$ | $*$ | $(7004)$ | 45 |
| 1991 | $*$ | 0 | 7 | 1 | 2 | 188 | 53 | $*$ | $*$ | $(250)$ | 58 |
| 1992 | 0 | 0 | 3 | 22 | 31 | 74 | 85 | 0 | 2 | 217 | 44 |
| 1993 | 0 | 0 | 0 | 0 | 0 | 24 | 8 | 87 | 4 | 122 | 69 |
| 1994 | 0 | 3 | 0 | 0 | 12 | 41 | 0 | 1 | 0 | 58 | 43 |
| 1995 | 0 | 0 | 0 | 3 | 2 | 158 | 22 | 2 | 5 | 190 | 67 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 16 | 26 | 21 | 49 | 112 | 41 |
| 1997 | 0 | 0 | 0 | 2 | 2 | 60 | 0 | 0 | 0 | 64 | 65 |
| 1998 | 0 | 0 | 0 | $<1$ | 0 | 0 | 55 | 57 | 4 | 117 | 43 |
| 1999 | 0 | 1 | 4 | 38 | 5 | $<1$ | 13 | 1 | 0 | 64 | 31 |
| 2000 | 0 | 63 | 65 | 80 | 60 | 27 | 6 | 2 | 56 | 360 | 20 |
| 2001 | 0 | 0 | 9 | 126 | 38 | 72 | 186 | 67 | 110 | 609 | 26 |
| 2002 | 0 | 0 | 9 | 59 | 96 | 52 | 1629 | 38 | 87 | 1967 | 48 |

Table 5.1.10 Cod off Greenland (offshore component), Greenland survey. Age disaggregate abundance indices (1000) for West Greenland, 1992-2002.

| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | $78+$ | TOTAL |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1992 | 0 | 221 | 126 | 123 | 63 | 10 | 3 | 1 | 547 |
| 1993 | 0 | 39 | 170 | 73 | 16 | 7 | 1 | 2 | 308 |
| 1994 | 0 | 10 | 126 | 22 | 8 | 1 | 0 | 0 | 167 |
| 1995 | 19 | 345 | 101 | 157 | 40 | 0 | 0 | 0 | 662 |
| 1996 | 0 | 14 | 203 | 78 | 3 | 0 | 0 | 0 | 298 |
| 1997 | 0 | 0 | 10 | 3 | 24 | 8 | 1 | 0 | 46 |
| 1998 | 0 | 17 | 25 | 20 | 0 | 0 | 0 | 0 | 62 |
| 1999 | 7 | 144 | 66 | 23 | 6 | 1 | 1 | 1 | 249 |
| 2000 | 90 | 711 | 363 | 92 | 13 | 52 | 0 | 0 | 1321 |
| 2001 | 97 | 540 | 546 | 376 | 0 | 0 | 0 | 0 | 1559 |
| 2002 | 0 | 603 | 2323 | 1078 | 245 | 0 | 4 | 0 | 4253 |



Figure 5.1.1

> Cod off Greenland. Catches 1955-2002 as used by the Working Group, inshore and offshore by West and East greenland (Horsted 1994,2000).


Figure 5.1.2 Cod off Greenland (offshore component), German survey. Survey area, stratification and position of hauls carried out in 2002.


Figure 5.1.3 Cod off Greenland (offshore component), German survey. Aggregated survey biomass indices for West and East Greenland and spawning stock biomass, 1982-2002. *)incomplete survey coverage.


Figure 5.1.4 Cod off Greenland (offshore component), Greenland survey. Aggregated survey biomass indices for West Greenland, 1992-2002.


Figure 5.1.5 Comparison of survey estimates of abundance at age 3 in a given year with age 1 two years earlier $\left(r^{2}=0.70\right)$ and with age 2 one year earlier $\left(r^{2}=0.90\right)$ for East and West Greenland offshore cod. Years with incomplete coverage off East Greenland omitted. Data derived from Working Doc 3/2003.


Figure 5.1.6 Weighted mean length-at-age 1-10 years 1982, 1984-2002 sampled in West Greenland (offshore component). Data derived from Working Doc 3/2003.


| - | age -1 |
| :---: | :---: |
| 0 | age -2 |
| $\nabla--$ | age -3 |
| $\nabla$ - - | age -4 |
| - | age -5 |
| - - - - | age -6 |
| $\square-$ | age -7 |
| $\checkmark$ | age -8 |
| $\triangle$ | age -9 |
| $--\triangle---$ | age -10 |

Figure 5.1.7 Weighted mean length-at-age 1-10 years 1982, 1984-2002 sampled in East Greenland (offshore component). Data derived from Working Doc 3/2003.


Figure 5.1.8 Number of cod /hour trawl off Greenland (offshore component), Greenland survey. Survey area, stratification and position of hauls carried out in 2002.


Figure 5.1.9


Figure 5.1.10 Greenland cod (offshore component). Trends in spawning stock biomass (SSB) and recruitment.


Figure 5.1.11 The relationship between the historic VPA abundance (4+) and survey abundance (4+) in the overlapping time period 1982-1993 (offshore component).


Figure 5.1.12 The relation between the abundance of the $4+$ group from the historic VPA and the German survey (converted to the VPA by the regression from figure 5.1.11) versus the 3 years old recruits. The very high recruit number ( 265 mil.) with corresponding low $4+$ group is the 1984-year class from Iceland. The high number of 4+ with corresponding very low recruits is also caused by the 1984 year class homing to Iceland before spawning (offshore component).

Spawning cod is documented for several fjords and costal areas between 64 and $67^{\circ} \mathrm{N}$ in West Greenland (Hansen 1949, Smidt 1979, Buch et al., 1994). The inshore cod populations are believed to be relatively stationary, as most (82-86\%) of the cod recaptured were found in the same area as they were tagged (Hovgård and Christensen 1990). Some interactions between the offshore and inshore cod stocks probably exist as the strong 1984- and partly 1985 year class was registered in the inshore gillnet survey as well as in the inshore landings. These strong year classes are believed to be Icelandic cod spawned off South-western Iceland. Some year's larvae are carried by the Irminger current to settle in South and West Greenland and contribute to the local fjord populations (Wieland and Hovgaard 2002).

### 5.2.1 Trends in Landings and Effort

The Greenland commercial cod fishery started locally in West Greenland in 1911 at some localities where cod seemed to occur regularly during summer and autumn. It took 15 years to reach 1000 (Hansen 1949). In 1924 an offshore fishery started and until 1974 the inshore catches have been of limited importance accounting for only $5-15 \%$ of the total fishery in Greenland water. Annual catches above 20000 t have been taken inshore during the period 1955-1969 and in 1980 and 1989 catches of approximately 40000 t were landed, partly driven by a few strong year classes entering from the offshore stock (Horsted 2000). Due to the very low offshore catches the importance of the inshore landings has increased accounting for between 50-90\% landings in the period 1993-2002. In the same period the inshore landings have been fluctuating between 500-4 000t.

In 1998 the lowest catch since 1918 was registered with at 326t. Slight improvements have been registered since 1999 with catches increasing to approximately 1700 t in 2001 and the preliminary statistics for the catches in 2002 is close to 4000 t. Especially NAFO division 1 B has experienced an increase, accounting for nearly $70 \%$ of the total inshore landings in 2001 (table 5.2.1).

Pound nets, gillnets and handlines are used to take about $95 \%$ of the inshore catch.
A commercial pound net CPUE series is available between 1992-1999. The mean catch per pound net setting decreased from 804 t in 1994 to 284 in 1999. No commercial effort data from 2000 to 2002 and catch-at-age data in 1997-1998 and 2000-2001 have been available to the working group.

### 5.2.2 West Greenland young cod survey

A survey using gangs of gillnets with different mesh-sizes has been conducted since 1985 with the objective to assess the abundance and distribution of pre-recruit cod in inshore areas of Greenland. The survey has usually been carried out in three inshore areas off West Greenland: Qaqortoq (NAFO Div. 1F), Nuuk (Div. 1D) and Sisimiut (Div. 1B). The Greenland inshore cod stock is not distributed in the Qaqortoq area, but occasional in $\mathbf{F}_{\text {low }}$ of pre-recruited cod from East Greenland and Iceland shows up here. Technical problems caused that only Division 1D was covered in 1999, and again in 2000 only Div. 1D and Div. 1F was covered. A more detailed description of the survey is provided in the 2001 report and WD 16/2003. No survey took place in 2001 and in 2002 Div. 1B and 1D were covered.

The recruitment index of 2-year old cod is shown in Figure 5.2.1 and reveals a strong 1984 -year class, a moderate 1985,1987, 1990- and 1993-year class and four successive weak year classes up to 2000. The very low 1997- and 1998class year might not be representative due to insufficient survey coverage. An increase in 2-year recruits was observed in 2002 Div 1B, reaching the levels from 1986-87 suggesting a strong 2000 year class in this division however as this area has not been covered during the three previous years, the size of the year class remains uncertain. In Div 1D the depleted status of the stock was confirmed in 2002 being only $7 \%$ of the 1984 year class.

### 5.2.3 Assessment of the stocks

Previously an Schaefer general production model was fitted to the Greenland inshore cod landing data using the commercial pound net CPUE results for 1993 to 1997 as an index of stock biomass. Lack of contrast in data impeded the model to run satisfactory.

Catch-at-age data for the period 1985-1996 and for 1999 and 2002 were available to the working group (Table 5.2.2). A statistical age structured model implemented MS Excel on the inshore cod stock was used as an exploratory tool to estimate the likely historical stock and exploitation dynamics. The model is based on a forward projection of stock in numbers, estimating initial stock size, selectivity by age, fishing mortality of each year and catch-at-age and minimizes the latter with the observed catch-at-age. A natural mortality of 0.2 was used as a scaling factor. The analysis was
stabilized by tuning it with age-based survey index for age groups 1 to 4 from the West Greenland young cod survey (Table 5.2.3). The selection pattern was assumed to be the same over the time period. Selectivity was estimated for age groups 2,3 and 4 but set to unity for older age groups. The relationship between population numbers and survey indices was assumed to be of a simple linear form (i.e. no power function applied). The error structure of the catches and the survey was assumed to be of lognormal form in the minimization function. The age classes where given different weights, the procedure being described in the WD32/2003. Equal weights were given to both input matrices and a penalty was added to force the predicted yield with that observed.

The residuals suggest that the data are relatively noisy (Figure 5.2.2). A positive trend in the survey data compared to the commercial data is observed from the beginning of the series until and including 1998. Applying a simple linear model to the estimator of survey catchability for the period 1985-1998 indicate an increase in survey catchability by about $22 \%$ per year. The apparent efficiency increase in the survey needs further investigation.

The historical stock dynamics indicate that the stock is currently around $10 \%$ of the maximum (Figure 5.2.3). Fishing mortality of the stock has been relatively low in recent years, but may be increasing. There are indication of improved recruitment relative to that observed in the last decade.

The residual pattern in the survey as well as lack of catch-at-age data in many of the recent years make the current assessment a relatively poor basis for mangement advice. Continous annual measurements and ageing of the catch composition as well as the continuation of the survey may however change that in the forseeable future.

### 5.2.4 Biological reference points

No specific values can be put forward as reference points due to the depleted state of the stocks.

### 5.2.5 Management Considerations

The inshore fishery exploiting possible self-sustained local fjord populations off West Greenland has historically been small, and the fishery has never been constricted by regulations. The data presented indicate that the stock has undergone a series of recruitment poor in recent years, but recovery potential is observed in Div 1B in 2002. No fishing should take place at this time until continues increase in recruitment and CPUE is evident.

Table 5.2.1 Cod catches divided to NAFO -divisions, caught inshore from vessels $>50$ GRT (Horsted 2000, Statistic Greenland 2002). *Not broken down to NAFO division.

| Year\Div | Nafo 1A | Nafo 1B | Nafo 1C | Nafo 1D | Nafo 1E | Nafo 1F | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 175 | 3908 | 1889 | 5414 | 1149 | 1333 | 19958 |
| 1985 | 149 | 2936 | 957 | 1976 | 1178 | 1245 | 8441 |
| 1986 | 76 | 1038 | 255 | 1209 | 1456 | 1268 | 5302 |
| 1987 | 97 | 2995 | 536 | 8110 | 4560 | 1678 | 8402 |
| 1988 | 333 | 6294 | 1342 | 2992 | 3346 | 4484 | 22829 |
| 1989 | 634 | 8491 | 5671 | 8212 | 10845 | 4676 | 28529 |
| 1990 | 476 | 9857 | 1482 | 9826 | 1917 | 5241 | 29026 |
| 1991 | 876 | 8641 | 917 | 2782 | 1089 | 4007 | 18311 |
| 1992 | 695 | 2710 | 563 | 1070 | 239 | 450 | 5723 |
| 1993 | 333 | 323 | 173 | 968 | 18 | 109 | 1924 |
| 1994 | 209 | 332 | 589 | 914 | 11 | 62 | 2115 |
| 1995 | 53 | 521 | 710 | 332 | 4 | 81 | 1710 |
| 1996 | 41 | 211 | 471 | 164 | 11 | 46 | 948 |
| 1997 | 18 | 446 | 198 | 99 | 13 | 130 | 1186 |
| 1998 | 9 | 118 | 79 | 78 | 0 | 38 | 319 |
| 1999 | 68 | 142 | 55 | 336 | 8 | 4 | 622 |
| 2000 | 154 | 266 | 0 | 332 | 0 | 12 | 764 |
| 2001 | 117 | 1183 | 245 | 54 | 0 | 81 | 1680 |
| 2002 |  |  |  |  |  |  | $3698^{*}$ |

Table 5.2.2 Catch-at-age (abundance in millions) 1985-2002, missing values in 1997,1998,2000 and 2001.

| Year\Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 |  |  |  | 0.742 | 0.588 | 2.464 | 0.154 | 0.604 | 0.016 |
| 1986 |  | 0.043 | 0.594 | 7.638 | 0.170 | 1.245 | 0.117 | 0.565 | 0.014 |
| 1987 |  | 0.052 | 0.214 | 7.533 | 6.446 | 0.320 | 0.877 | 0.229 | 0.415 |
| 1988 | 0.006 | 0.218 | 11.813 | 12.619 | 1.318 | 0.452 | 0.369 | 0.172 | 0.184 |
| 1989 |  | 0.002 | 0.154 | 10.169 | 9.340 | 2.632 | 0.742 | 0.137 | 0.116 |
| 1990 | 0.004 | 0.125 | 7.177 | 8.562 | 2.499 | 0.288 | 0.012 | 0.003 |  |
| 1991 |  | 0.001 | 0.051 | 1.767 | 2.634 | 0.730 | 0.126 | 0.008 | 0.005 |
| 1992 | 0.000 | 0.029 | 0.647 | 0.706 | 0.208 | 0.044 | 0.006 | 0.006 |  |
| 1993 |  | 0.001 | 0.053 | 1.152 | 0.727 | 0.079 | 0.053 | 0.012 | 0.003 |
| 1994 |  | 0.008 | 0.593 | 0.729 | 0.140 | 0.036 | 0.001 | 0.001 |  |
| 1995 |  | 0.002 | 0.148 | 0.262 | 0.119 | 0.056 | 0.009 | 0.007 |  |
| 1996 |  |  |  |  |  |  |  |  |  |
| 1997 |  |  |  |  |  |  |  |  |  |
| 1998 |  |  |  |  |  |  |  |  |  |
| 1999 |  |  | 0.082 | 0.396 | 0.238 | 0.037 | 0.004 |  |  |
| 2000 |  |  |  |  |  |  |  |  |  |
| 2001 |  |  |  |  |  |  |  |  |  |
| 2002 |  |  |  |  |  |  |  |  |  |

Table 5.2.3 CPUE (number of age 1,2,3 and 4 cod caught per 100 hours net setting) in the Greenland Gillnet cod survey covering West Greenland 1987-2002.

| Age | 1 | 2 | 3 | 4 |
| :---: | :---: | :---: | :---: | :---: |
| 1985 | 107.51 | 45.36 | 0.37 | 2.53 |
| 1986 | 6.22 | 124.04 | 11.77 | 1.26 |
| 1987 | 0.34 | 75.04 | 119.82 | 6.73 |
| 1988 | 0.03 | 15.27 | 72.32 | 34.32 |
| 1989 | 0.11 | 58.47 | 37.33 | 21.67 |
| 1990 | 0.00 | 24.12 | 34.95 | 12.22 |
| 1991 | 63.63 | 2.40 | 29.00 | 12.16 |
| 1992 | 0.10 | 38.22 | 13.14 | 7.69 |
| 1993 | 0.00 | 6.89 | 33.20 | 10.45 |
| 1994 | 0.65 | 1.40 | 6.37 | 4.32 |
| 1995 | 0.23 | 18.95 | 3.76 | 3.16 |
| 1996 | 0.00 | 7.45 | 10.32 | 1.66 |
| 1997 | 1.92 | 5.88 | 2.71 | 0.82 |
| 1998 | 0.32 | 7.66 | 13.46 | 1.28 |
| 1999 | 0.00 | 0.40 | 1.20 | 2.70 |
| 2000 | 0.12 | 6.96 | 4.14 | 0.40 |
| 2001 | no | survey |  |  |
| 2002 | 7.25 | 53.24 | 19.61 | 6.89 |



Figure 5.2.1 CPUE (number of age 2 cod caught per 100 hours net setting) in the Greenland Young cod survey 1985-2002 (inshore component). The three areas covered in the survey are shown in triangel (NAFO 1 B), squares (NAFO 1D) and diamonds (NAFO 1F).



Figure 5.2.2
Greenland cod (inshore component). Catch-at-age model showing the residuals in the two input components survey and catch-at-age matrix.


Figure 5.2.3 Greenland cod 1985-2002 (inshore component). Catch-at-age model showing fishing mortality, exploitation rate the residuals between the two input components survey and catch-at-age matrix.

### 6.1 Landings, Fisheries, Fleet and Stock Perception

Total annual landings in Divisions Va, Vb, and Subareas XII and XIV are presented for the years 1981-2002 in Tables 6.1.1-6.1.5 and since 1961 in Figure 6.1.1. Landings during the decade prior to the extension of the EEZ to 200 nm by coastal nations in 1976 were in the order of 20-35 kt. From 1976, landings increased from a low of 5 kt to above 30 kt after 1982. In the years 1987-1989, landings increased to about 61 kt , followed by a decrease to about 35-40 kt during 1992-96. After 1996, landings declined to 20 kt in 1998 and 1999. Since 2000 an increase in landings has been observed, to nearly 26 kt in 2000 and about 28 kt in 2001 and 29 kt in 2002. Landings not officially reported to ICES have been included in the assessment.

Catches in Icelandic waters have, due to quota regulations, decreased from 37 kt in 1990 to 11 kt in 1998 and 1999, but have risen again to 14 kt in 2000 and about 16 kt in 2001 and 19 kt in 2002 Faroese catches in Vb increased from 1 kt in 1981-1991 to 6.5 kt in 1996, but was of the order of $4-5 \mathrm{kt}$ in during 1997-2001. In 2002 catches decreased to 2.6 kt . Catches in Division XIVb have increased from below 1 kt in 1987-1991 to 8.5 kt in 1997, followed by a decrease to 5 kt in 1999. Since then catches have increased to about 7 kt .

Most of the fishery for Greenland halibut in Divisions $\mathrm{Va}, \mathrm{Vb}$ and XIVb is a directed fishery, only minor catches in Va by Iceland, and the catches in XIVb by Germany and the UK are by-catches in redfish fisheries. A detailed description of the fishery performance and areas is given in NWWG report 1998. No major changes were observed in 2002. Table 6.1.6 describes the Working Group's best landing estimates for the year 2002 with respect to area and gear. In the Greenland EEZ, only about $50 \%$ of the national allocated total quotas were fished.

## Stock perception

The current definition of the Greenland halibut in East Greenland, Iceland, and Faroe waters as one stock, specified by ICES in 1976 was "based on a strong probability that the spawning grounds [for Greenland halibut in these waters] are the same". A summary of the current state of knowledge on Greenland halibut in the above-mentioned waters shows that key information on the life cycle is lacking (Woll 2000). Information on the spawning location and spawning time of the stock is very limited. It is hypothesised, based on information from one scientific bottom trawl cruise in 1977, that the major spawning grounds are located on the continental slopes west of Iceland at depths around and below 1000 m (Magnusson 1977; Sigurdsson 1977; Sigurdsson and Magnusson 1980). In recent years (1995 and 2000), some spawning has been observed in East Greenland waters ( $62^{\circ} \mathrm{N}$ and $64^{\circ} \mathrm{N}$ ) in August (Gundersen et al. 1997; Fossen and Gundersen 2000).

Standard 0-group fish surveys have been carried out annually in late summer (mainly in August) in Icelandic and in East Greenland waters since 1970. Larvae are mainly observed along the shelf region off East Greenland and are in some years abundant all over the shelf area south to $60^{\circ} \mathrm{N}$, which is the southernmost limit of the survey area. Highest abundance is observed on the continental shelf north of $64^{\circ} \mathrm{N}$ and just east off the continental shelf south of $64^{\circ} \mathrm{N} .0-$ group larvae are only occasionally observed on the Icelandic shelf in very limited numbers. Nursery grounds for young Greenland halibut (ages 1-3, fish less than 45 cm long) are well known in West Greenland waters, where they are most abundant from Store Hellefiske Bank to Disko and in Disko Bay between $66^{\circ}-69^{\circ}$ latitude at depths of about 200 m (Riget and Boje, 1988). When it comes to knowledge on young fish in East Greenland and Icelandic waters, information is very sparse. A gillnet survey targeting young Greenland halibut, modelling of advection of eggs and larvae with currents from assumed spawning areas in Icelandic and East Greenland waters (Woll 2000), and results of historic Greenland ichtyoplankton surveys (Boje 1997), indicated that larvae were transported to Southwest Greenland waters before settling, mixing with specimens from the Greenland-Canadian stock complex. Analyses of shrimp surveys in Icelandic and Greenland waters (Boje and Hjørleifsson 2000) concluded that nursery grounds were neither to be found in Icelandic nor in East Greenland waters.

The highest aggregation of commercial-sized Greenland halibut is found just south of the Greenland-Iceland ridge. In this area the major portion of the annual catch in the past 10 to 15 years has been taken mainly at depths between 500 and 1000 meters. Other locations of Greenland halibut in exploitable densities (for trawl fisheries) are found along the north and east coast of Iceland, mainly at depths between 500 to 700 meters, in waters of Faroe Islands, as well as along the continental slope off East Greenland. The sizes of the Greenland halibut in the trawl fisheries depend largely on location and depth, and to some extent on the season. In Icelandic waters, smaller fish are found along the east and north coast, with somewhat larger fish in the deeper waters south of the Faroe-Iceland ridge. The largest fish are, however, always found on the main fishing grounds between Iceland and Greenland.

Indices of CPUE for the Icelandic trawl fleet for the period 1985-2002 (Table 6.2.1, Fig. 6.2.1) were estimated from a GLIM multiplicative model, taking into account changes in the Icelandic trawl catch due to vessel, statistical square, month, and year effects. All hauls with Greenland halibut exceeding $50 \%$ of the total catch were included in the CPUE estimation. The CPUE indices from the trawling fleets in Divisions Va, Vb and XIVb were used to estimate the total effort for each year ( $y$ ) for each of the divisions according to:

$$
E_{y, d i v}=Y_{y, d i v} / C P U E_{y, d i v}
$$

where E is the total effort and Y is the total reported landings (Table 6.2.1).
Catch rates of Icelandic bottom trawlers decreased for all fishing grounds during 1990-1995, but stabilised in 19951997. In 1998, an increase of $60 \%$ in CPUE was observed for all fishing grounds coinciding with a drastic ( $60 \%$ ) reduction in effort (Table 6.2.1, Figure 6.2.1). In 1999 to 2001 CPUE increases annually between $4-15 \%$ until 2002 when CPUE decreased by $24 \%$. The total effort increased up to 1995, decreased significantly until 2001, but increased again in 2002 by $54 \%$. Effort during 1998-2001 has been less than half of that in 1995-97.

Information from logbooks from the Faroese otterboard trawl fleet ( $>1000 \mathrm{hp}$ ) was available for the years 1991-2002 (Table 6.2.1), which represents an extension in the time-series back in time as compared to last year (1995-2001). It is a rather new fishery and the location of the fishery has changed from the eastern side of the islands in 1995-1998, to the western side in 2000 . Therefore, the fishery is assumed to be in the process of learning. Only hauls where G.halibut consisted of more than $50 \%$ of the catches and conducted on depths more than 450 meters were selected for the analyses. The logbooks were standardised with a multiplicative model using $\operatorname{logCPUE}$ as dependent variable, taking into account locality, vessel, month, and year. The fishery is fairly new in the area and has increased from about 1500 t in 1991 to 5000 t in 2000. CPUE decreased in the early period by about $10 \%$ coinciding with a significant increase in effort. Since 1994 CPUE's have been stable and effort has thus followed the development of the catches (Fig. 6.2.1).

For Division XIVb, logbook data was available from German, Norwegian, Faroese, Russian, Japanese and Greenland fleets. Hauls where targeted species was G.halibut and where catch weight exceeds 100 kg were selected as no information on other species caught was available. CPUE from logbooks in the years 1991-2002 were standardised using a multiplicative model taking into account locality, fleet, month and year and logCPUE as dependent variable (Table 6.2.1, Fig.6.2.1). CPUE increased significantly from 1993 to 1994, where after it remains relatively stable. Effort increased continuously until 1997, but declined by $30 \%$ until 1999. Since 1999 the effort increased and is high in 2002. However, the fishery in XIVb is just starting out and catches have increased from below 500 tons annually before 1991 to about 7000 t in the last three years. The fishery was therefore assumed to be in the process of learning in the beginning of the CPUE series, and subsequently the increase in CPUE should not be taken indicative for the stock development.

The three CPUE series from Divisions $\mathrm{Va}, \mathrm{Vb}$ and XIVb show contradictory trends in the period 1991 to 2002 (Fig.6.2.1). CPUE's in Vb and XIVb area stable for the period 1994 to 2002, while those series shows contradicting trends prior to 1994. In XIVb CPUE's increased from 1993 to 1994, while CPUE's decreased for Div. Vb in the same period. The Icelandic CPUE's (Va) shows yet another trend, decreasing since the late 1980'ies until 1996. From 1996 to 2001 CPUE's increased somewhat but decline again in 2002. This could indicate different stock status in the areas, but could also be artefacts, i.e. due to different behaviour of the fleets, migration between areas or/and caused by the different criteria for selection of hauls included in the standardisation. A compilation of an overall database of logbook data from all the areas is undertaken.

### 6.3 Catch-at-age

Age-length keys for 2002 were from: the Icelandic trawl fleet operating in Icelandic waters ( 424 otoliths). This key was used to obtain catch in number for the length samples for each of the following commercial fleets and areas:

| Gear | Area | Landings | No. samples | No. fish | A/L-Key |
| :--- | :--- | ---: | ---: | ---: | ---: |
| Bottom trawl | Iceland-west | 16337 | 111 | 14571 | Icelandic bottom trawl |
| Bottom trawl | Iceland-north \& east | 1476 | 13 | 1199 | Icelandic bottom trawl |
| Bottom trawl | Iceland-southeast | 1901 | 14 | 1665 | Icelandic bottom trawl |
| Gill Net (\&line) | Faroe Islands | 1975 |  | 1421 | Icelandic bottom trawl |
| Bottom trawl | Faroe Islands | 821 |  | 449 | Icelandic bottom trawl |
| Long line | East Greenland | 795 | 73 | 6702 | Icelandic bottom trawl |
| Bottom trawl | East Greenland | 3797 | 19 | 1953 | Icelandic bottom trawl |
| Total |  | 27102 | 230 | 27960 |  |

The following length-weight relationships were applied to convert sampled lengths to weights:

| Gear | Area | Length - weight key | Comments |
| :--- | :--- | :--- | :--- |
| Bottom trawl | Iceland | $\mathrm{W}=0.01758 \mathrm{~L}^{2.84387}$ | Same key as 1999 |
| Bottom trawl | East Greenland | $\mathrm{W}=0.00161^{*} \mathrm{~L}^{3.4457}$ | Commercial trawl $\mathrm{N}=2468$, same as 2001 |
| Bottom trawl | Faroe Islands | $\mathrm{W}=0.00202 \mathrm{~L}^{3.398}$ | Trawl survey, $\mathrm{N}=1916$ |
| Gill Net (\&line) | Faroe Islands | $\mathrm{W}=0.00202 * \mathrm{~L}^{3.398}$ | Trawl survey, $\mathrm{N}=1916$ |
| Long line | Iceland | $\mathrm{W}=0.01758 \mathrm{~L}^{2.84387}$ | Same key as 1999 |
| Long line | East Greenland | $\mathrm{W}=0.000041^{*} \mathrm{~L}^{3.245}$ | Longline data, 2002, $\mathrm{N}=3175$ |

The total catch in numbers (Table 6.3.1) was obtained from the sum of the above weighted with the catch within each group. Data for 1994 and 1996 - 2002 derives from several nations fishing, while only Icelandic data has been available for the remaining years back in time.

### 6.4 Weight-at-age

The mean weight-at-age in 2002 (Table 6.4.1) was derived from the weighted average of the above groups. Weights-atage in the catch are also used as weights-at-age in the stock. The suspect weight-at-age pattern as observed in 2002 is due to a change in the age-length key (age-readings). As growth of Greenland halibut is known not to show such variability, the age-readings will be further inspected before included in any age based assessment.

### 6.5 Maturity-at-age

Maturity data were not updated for 2002 as visual determination of maturity has been questionable as stated in 2001 report. Maturity for 2002 is thus compiled as an average of 1996-2000.

### 6.6 Survey information

An October groundfish survey in Icelandic waters, covering the distributional area of Greenland halibut within the Icelandic EEZ, was started in 1996. The survey is a fixed station stratified random survey consisting of 300 stations on the continental shelf and slope down to a depth of 1300 m . An increase in the fishable biomass of Greenland halibut (fish of length equal to or greater than 50 cm ) is observed from 1996 to 2001 (Figure 6.6.1b). Abundance indices of smaller fish ( $<50 \mathrm{~cm}$ ) indicate signs of improved recruitment in 1998 and 1999 that may account for the increase in the estimated fishable biomass over the period. In 2002 abundance of all length groups are below that observed in 2001. Biomass indices have increased in the survey since the beginning in 1996, although variable since 1999 (Fig. 6.6.1a).

Since 1998, a Greenland survey for Greenland halibut has been carried out in East Greenland waters from $60^{\circ} \mathrm{N}$ to $67^{\circ} \mathrm{N}$ at the main commercial fishing grounds at depths of $400-1500 \mathrm{~m}$ in late June/early July. In 2002 a total of 40 stations were hauled. No survey took place in 2001. Total estimated biomass in 2002 was estimated at 15 kt , which is a $40 \%$ decrease (not significant at the $95 \%$ level) from the 2000 biomass estimate (Fig. 6.6.2). The age composition in the survey does indicate a decrease in abundance of juveniles for 2002.

### 6.7.1 Age-based assessement

Age-disaggregated CPUE values for age groups 7-12 from the Icelandic trawling fleet operating in Division Va have previously been used in the XSA tuning assessments. Since 2000 the XSA assessment has been considered unreliable due to poor diagnostics mainly caused by inconsistent sampling and age readings (see section 6.9), and was thus rejected as a basis for advice. No attempt was made this year to run an age-based assessment due to the questionable input data. In the 2002 report is given the historic trends in $\log (q)$ residuals and the retrospective pattern of $F$. Based on those plots the Working Group in 2002 decided that an XSA model was not a reliable estimator of recent stock history.

### 6.7.2 Stock production model

A stock-production model approach, ASPIC, was attempted on the various indices and catches. ASPIC requires series of catch data and indices of stock biomass, either corresponding effort, CPUE, or survey catch rates. Corresponding catch and effort data is available for Div. Va, (formerly used as a tuning fleet in the XSA), Vb and XIVb, and in addition several survey series (Figure 6.2.1) were available:

| Fleet and index | Period | Division |
| :--- | :--- | :--- |
| Icelandic trawler CPUE from GLIM | $1973-2002$ | Va |
| Icelandic fall groundfish survey | $1996-2002$ | Va |
| Icelandic shrimp fishery | $1986-1994$ | Va |
| Icelandic shrimp survey | $1987-2000$ | Va |
| Greenland trawler CPUE from GLIM | $1991-2002$ | XIVb |
| Greenland spring deepwater bottom-trawl survey | $1997-2000,2002$ | XIVb |
| Faroese trawler CPUE from GLIM | $1991-2002$ | Vb |

The Icelandic shrimp fishery no longer exploits Greenland halibut, because of implementation of sorting grids in recent years. It does thus not provide indices of recent stock trends and was thus not included in the model. Since the shrimp survey covers a relatively limited area, the index was also excluded as an input candidate into the model. The Greenland deepwater survey only consist of a short time-series with lack of a 2001 survey and was therefore not used. A run using the remaining four indices failed due to conflicting trends for the CPUE series in Divs. XIVb and Vb in the early 1990'ies. For the two remaining indices - Icelandic trawler standardized CPUE and Icelandic groundfish survey ASPIC was run with a reduced commercial time-series from 1985-2001 and the fall groundfish survey from 1996-2001. The decision of using only a reduced time-series is because the CPUE index from 1973 to 1985 may not be reliable because it is based on limited logbook material and may cover a learning period at the beginning of the fishery.

ASPIC (BETA vers 4.45) requires starting guesses for $K$, carrying capacity, MSY and $\mathrm{B} 1 / \mathrm{K}$ ratio (Initial biomass $/ \mathrm{K}$ ). ASPIC was run fitting a logistic model conditioned on catch as in previous two years. Initially ASPIC was run with different starting guesses of these parameters to explore stability of parameter estimation. For an appropriate range of input values, ASPIC results were incredibly stable. The parameter estimates from ASPIC are comparable to last year (Table 6.7.2.1.). MSY is estimated to 35 kt and $\mathbf{B}_{\text {MSY }}$ to 114 kt . Biomass in 2003 is estimated to be about $22 \%$ below $\mathbf{B}_{\text {MSY }}$ and fishing mortality in 2002 is estimated to be $10 \%$ above $\mathbf{F}_{\text {MSY }}$. Observed and estimated CPUE's are provided in Fig. 6.7.2.1.

The state of the stock relative to $\mathbf{F}_{\mathrm{MSY}}$ and $\mathbf{B}_{\mathrm{MSY}}$ is given in the Fig 6.7.2.2. Biomass is increasing from a record low in 1998 and in 2003 is about $25 \%$ below $\mathbf{B}_{\text {MSY }}$. F has in the last decade been very high ( $60 \%$ above $\mathbf{F}_{\text {MSY }}$ ), but since 1998 is estimated to be near or above $\mathbf{F}_{\text {MSY }}$.

Retrospective analyses were carried out in the 2002 NWWG report for both $\mathrm{B} / \mathbf{B}_{\text {MSY }}$ and $\mathrm{F} / \mathbf{F}_{\text {MSY }}$ in order to exploit the consistency of ASPIC with the currently used CPUE series. ASPIC behaves consistent when contrasting data is available, e.g. back to about 1997.

### 6.7.3 Stock projection

From calculated stock-dynamic parameters and input fishing regimes, ASPIC can project forward trajectories of population biomass and fishing mortality including uncertainty estimates based on bootstrapping. In all forward projections it was assumed that the catch in 2003 would be maintained at 30 kt . This is based on the following: TAC in

Icelandic waters is maintained at 20 kt and expected to be caught. Given that the landings in Vb and XIV will be the same as in 2002 and that the Icelandic fleet will catch all its quota, it is anticipated that total landings in the year 2003 also will be in the order of 30 kt . Three different trajectories were produced using the following options:

1) $\mathrm{F}(2002-10)=2 / 3 \mathbf{F}_{\mathrm{MSY}} \sim \mathbf{F}_{\mathrm{pa}}$,
2) $F(2002-10)=\mathbf{F}_{\text {sq }}$,
3) $\operatorname{Catch}(2001-2010)=30000 \mathrm{t}$.

Plots of B-ratios $\left(\mathrm{B} / \mathbf{B}_{\mathrm{MSY}}\right)$ are given in Figure 6.7.3 and biomass trajectory for option 1 only is given in Table 6.7.3. By fishing at $\mathbf{F}_{\mathrm{pa}}\left(2 / 3 \mathbf{F}_{\mathrm{MSY}}\right)$ it is expected that the biomass will increase above $\mathbf{B}_{\mathrm{MSY}}$ by 2006 . Fishing at $\mathbf{F}_{\mathrm{sq}}$ will result in $\mathbf{B}_{\mathrm{MSY}}$ never be achieved, although the $80 \%$ confidence interval includes attaining $\mathbf{B}_{\mathrm{MSY}}$. Fishing at 30 kt annually is expected to allow recovery to $\mathbf{B}_{\mathrm{MSY}}$ by 2009 , but with a significant risk ( $80 \%$ confidence intervals) that the stock will collapse. Landings in 2004 associated with the trajectories are 20000 t at $\mathbf{F}_{\mathrm{pa}}$ and 34000 t at $\mathbf{F}_{\mathrm{sq}}$.

### 6.7.4 Biological reference points

Defined reference points for Greenland halibut have previously been defined on the basis of an age-based analytical assessment. The Working Group considers it appropriate to define $\mathbf{F}_{\mathrm{pa}}$ as $2 / 3$ of $\mathbf{F}_{\mathrm{MSY}}$ estimated from the stockproduction model. Using $2 / 3$ as $\mathbf{F}_{\mathrm{pa}}, \mathbf{F}_{\text {lim }}$ could be calculated using $\mathbf{F}_{\text {lim }}=\mathbf{F}_{\mathrm{pa}} * \mathrm{e}^{1.645 \sigma}$, where $\sigma$ could be 0.30 .

### 6.8 Management Considerations

No formal agreement on the management of the Greenland halibut exists among the three coastal states, Greenland, Iceland, and the Faroe Islands. The regulation schemes of those states have previously resulted in catches well in excess of TAC's advised by ICES. A likely scenario is therefore a continuation of status quo catch at 30 kt in the short-term. This will most probably result in a steady recovery of the biomass in the near future, but with a risk that the stock will remain low or even collapse.

### 6.9 Comments on the Assessment

An analytical assessment (XSA) was previously attempted but rejected due to poor diagnostics and a substantial new perception of the stock size. Both former XSA and the recent stock production model suggest that the Greenland halibut stock biomass has been falling since the late 1980'ies. Also according to both assessment methods, the fishing mortality has been substantially above $\mathbf{F}_{\mathrm{pa}}$ for a decade. The decline in biomass seems to have been halted since 1998, but biomass is still well below $\mathbf{B}_{\text {MSY }}$. A combination of unreliable maturity data and age readings from recent years still impede any age-disaggregated assessment, and therefore also any estimate of SSB and its use in relation to $\mathbf{B}_{\mathrm{pa}}$ and SSB as a reference point for management advice for the stock.

The stock production model used to assess the status of the stock relies on the same trawler CPUE series as previously used in the XSA. Output estimates of biomass and fishing mortality of the production model cannot be taken at face value, but should rather be good estimates of the state of the stock in relation to MSY parameters.

Use of other indices than the currently Icelandic (Va) CPUE series and survey series in the stock production model (ASPIC) should be explored. CPUE series from XIVb and Vb are presently available, but due to different trends than the Icelandic series it impede inclusion in the ASPIC.

Table 6.1.1. GREENLAND HALIBUT. Nominal catches (tonnes) by countries, in Sub-areas V, XII and XIV 1981-2002, as officially reported to ICES.

| Country | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | - | - | - | - | - | - | 6 | + | - |
| Faroe Islands | 767 | 1532 | 1146 | 2502 | 1052 | 853 | 1096 | 1378 | 2319 |
| France | 8 | 27 | 236 | 489 | 845 | 52 | 19 | 25 | - |
| Germany | 3007 | 2581 | 1142 | 936 | 863 | 858 | 565 | 637 | 493 |
| Greenland | + | 1 | 5 | 15 | 81 | 177 | 154 | 37 | 11 |
| Iceland | 15457 | 28300 | 28360 | 30080 | 29231 | 31044 | 44780 | 49040 | 58330 |
| Norway | - | - | 2 | 2 | 3 | + | 2 | 1 | 3 |
| Russia | - | - | - | - | - | - | - | - | - |
| UK (Engl. and Wales) | - | - | - | - | - | - | - | - | - |
| UK (Scotland) | - | - | - | - | - | - | - | - | - |
| United Kingdom | - | - | - | - | - | - | - | - | - |
| Total | 19239 | 32441 | 30891 | 34024 | 32075 | 32984 | 46622 | 51118 | 61156 |
| Working Group estimate | - | - | - | - | - | - | - | - | 61396 |


| Country | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | - | - | - | - | - | - | 1 | - |  |
| Faroe Islands | 1803 | 1566 | 2128 | 4405 | 6241 | 3763 | 6148 | 4971 | 3817 |
| France | - | - | 3 | 2 | - | - | 29 | 11 | 8 |
| Germany | 336 | 303 | 382 | 415 | 648 | 811 | 3368 | 3342 | 3056 |
| Greenland | 40 | 66 | 437 | 288 | 867 | 533 | 1162 | 1129 | 747 |
| Iceland | 36557 | 34883 | 31955 | 33987 | 27778 | 27383 | 22055 | 18569 | 10728 |
| Norway | 50 | 34 | 221 | 846 | $1173{ }^{1}$ | 1810 | 2164 | 1939 | 1367 |
| Russia | - | - | 5 | - | - | 10 | 424 | 37 | 52 |
| UK (Engl. and Wales) | 27 | 38 | 109 | 811 | 513 | 1436 | 386 | 218 | 190 |
| UK (Scotland) | - | - | 19 | 26 | 84 | 232 | 25 | 26 | 43 |
| United Kingdom |  |  |  |  |  |  |  |  |  |
| Total | 38813 | 36890 | 35259 | 40780 | 37305 | 36006 | 35762 | 30242 | 20360 |
| Working Group estimate | 39326 | 37950 | 35423 | 40817 | 36958 | 36300 | 35825 | 30267 | - |


| Country | $1999^{1}$ | $2000^{1}$ | 2001 | 2002 |
| :--- | :---: | :---: | ---: | ---: |
| Denmark |  | - | 0 | 0 |
| Faroe Islands | 3884 | - | 0 | 0 |
| France | - | 21 | 25 | 20 |
| Germany | 3082 | 3271 | 2807 | 2148 |
| Greenland | 200 | 1740 | 1553 | 0 |
| Iceland | 11180 | 14537 | 16590 | 19223 |
| Ireland |  | - | 7 |  |
| Norway | 1187 | 1272 | 1483 | 1328 |
| Portugal |  |  | 6 |  |
| Russia | 138 | 183 | 186 | 44 |
| Spain |  | 8 | 10 |  |
| UK (Engl. and Wales) | 261 | 370 | 227 |  |
| UK (Scotland) | 69 | 121 | 130 |  |
| United Kingdom | - | - |  | 441 |
| Total | 20001 | 21523 | 23024 | 23204 |
| Working Group estimate | 20371 | 26839 | 28021 | 29260 |

1) Provisional data

Table 6.1.2. GREENLAND HALIBUT. Nominal catches (tonnes) by countries, in Division Va 1981-2002, as officially reported to ICES.

| Country | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Faroe Islands | 325 | 669 | 33 | 46 |  |  | 1989 |  |
| Germany <br> Greenland |  |  |  |  |  |  |  |  |
| Iceland | 15455 | 28300 | 28359 | 30078 | 29195 | 31027 | 44644 | 49000 |
| Norway |  |  | + | + | 5830 |  |  |  |
| Total | 15780 | 28969 | 28392 | 30124 | 29197 | 31027 | 44659 | 49379 |
| Working Group estimate |  |  |  |  |  | 59049 |  |  |


| Country | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | 739 | 273 | 23 | 166 | 910 | 13 | 14 | 26 | 6 |
| Germany |  |  |  |  | 1 | 2 | 4 |  | 9 |
| Greenland |  |  |  |  | 1 |  |  |  | 1 |
| Iceland | 36557 | 34883 | 31955 | 33968 | 27696 | 27376 | 22055 | 16766 | 10580 |
| Norway |  |  |  |  |  |  |  | 1 | 1 |
| Total | 37296 | 35156 | 31978 | 34134 | 28608 | 27391 | 22073 | 16792 | 10595 |
| Working Group estimate | $37308^{2}$ | $35413{ }^{2}$ |  |  |  |  |  |  |  |


| Country | 1999 | 2000 | 2001 | $2002^{1}$ |
| :--- | ---: | ---: | ---: | ---: |
| Faroe Islands | 9 |  |  |  |
| Germany | 13 | 22 | 50 | 31 |
| Greenland | 1 |  |  |  |
| Iceland | 11087 | 14507 | 2310 | 4 |
| Norway |  |  | 6 |  |
| UK (E/W/I) | 26 | 73 | 50 |  |
| UK Scottland | 3 | 5 | 12 |  |
| UK |  |  |  | 37 |
| Total | 11138 | 14607 | 2428 | 19291 |
| Working Group estimate |  | $14519^{3}$ | 16752 | 19714 |

1) Provisional data
2) Includes $223 t$ catch by Norway.
3) Includes $12 t$ catch by Norway.
4) 14280 t fished in Icelandic EEZ, previously reported in Va, are in 2002 moved to ICES XIV b.

Table 6.1.3 GREENLAND HALIBUT. Nominal catches (tonnes) by countries, in Division Vb 1981-2002, as officially reported to ICES.

| Country | 1981 | 1982 | 1983 |  | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | - | - | - |  | - | - | - | 6 | + | - |
| Faroe Islands | 442 | 863 | 1112 |  | 2456 | 1052 | 775 | 907 | 901 | 1513 |
| France | 8 | 27 | 236 |  | 489 | 845 | 52 | 19 | 25 | $\ldots$ |
| Germany | 114 | 142 | 86 |  | 118 | 227 | 113 | 109 | 42 | 73 |
| Greenland | - | - | - |  | - | - | - | - | - | - |
| Norway | 2 | + | 2 |  | 2 | 2 | + | 2 | 1 | 3 |
| UK (Engl. and Wales) | - | - | - |  | - | - | - | - |  | - |
| UK (Scotland) | - | - | - |  | - | - | - | - |  | - |
| United Kingdom | - | - | - |  | - | - | - | - | - | - |
| Total | 566 | 1032 | 1436 |  | 3065 | 2126 | 940 | 1043 | 969 | 1589 |
| Working Group estimate | - | - | - |  | - | - | - | - | - | $1606^{2}$ |
| Country | 1990 | 1991 | 1992 |  | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| Denmark | - | - | - |  | - | - | - | - | - |  |
| Faroe Islands | 1064 | 1293 | 2105 |  | 4058 | 5163 | 3603 | 6004 | 4750 | 3660 |
| France 6 | ... | ... | 3 | 1 | 2 | 1 | 28 | 29 | 11 | $8^{1}$ |
| Germany | 43 | 24 | 71 |  | 24 | 8 | 1 | 21 | 41 |  |
| Greenland | - | - | - |  | - | - | - | - | - |  |
| Norway | 42 | 16 | 25 |  | 335 | 53 | 142 | 281 | 42 | $114{ }^{1}$ |
| UK (Engl. and Wales) | - | - | 1 |  | 15 | - | 31 | 122 |  |  |
| UK (Scotland) | - | - | 1 |  | - | - | 27 | 12 | 26 | 43 |
| United Kingdom | - | - | - |  | - | - |  |  |  |  |
| Total | 1149 | 1333 | 2206 |  | 4434 | 5225 | 3832 | 6469 | 4870 | 3825 |
| Working Group estimate | $1282{ }^{2}$ | $1662{ }^{2}$ | 2269 | 2 | - | - |  | - | - | 0 |


| Country | 1999 |  | 2000 | 1 | 2001 |  | 2002 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |  |  |  |  |
| Faroe Islands | 3873 |  |  |  |  |  |  |  |
| France |  |  | 21 |  | 25 | 1 | 20 |  |
| Germany | 22 |  | 6 |  | 7 |  |  |  |
| Iceland |  |  |  |  |  |  |  |  |
| Ireland |  |  |  |  | + |  |  |  |
| Norway | 87 |  | 110 | 1 | 53 | 1 | 48 |  |
| UK (Engl. and Wales) | 9 |  | 35 |  | 77 |  |  |  |
| UK (Scotland) | 66 |  | 116 |  | 118 |  |  |  |
| United Kingdom |  |  |  |  |  |  | 202 |  |
| Total | 4057 |  | 288 |  | 280 | 2 | 270 |  |
| Working Group estimate | 2694 | 2 | 5092 | 3 | 3951 |  | 2694 |  |

1) Provisional data
2) WG estimate includes additional catches as described in Working Group reports for each year and in the report from 2001.

Table 6.1.4 GREENLAND HALIBUT. Nominal catches (tonnes) by countries,
in Sub-area XIV 1981-2002, as officially reported to ICES.

| Country | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | - | - | - | - | - | 78 | 74 | 98 | 87 |
| Germany | 2893 | 2439 | 1054 | 818 | 636 | 745 | 456 | 595 | 420 |
| Greenland | + | 1 | 5 | 15 | 81 | 177 | 154 | 37 | 11 |
| Iceland | - | - | 1 | 2 | 36 | 17 | 136 | 40 | + |
| Norway | - | - | - | + | - | - | - | - | - |
| Russia | - | - | - | - | - | - | - | - | + |
| UK (Engl. and Wales) | - | - | - | - | - | - | - | - | - |
| UK (Scotland) | - | - | - | - | - | - | - | - | - |
| United Kingdom | - | - | - | - | - | - | - | - | - |
| Total | 2893 | 2440 | 1060 | 835 | 753 | 1017 | 820 | 770 | 518 |
| Working Group estimate | - | - | - | - | - | - | - | - | - |
|  |  |  |  |  |  |  |  |  |  |
| Country | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| Denmark | - | - | - | - | - | - | 1 | + | + |
| Faroe Islands | - | - | - | 181 | 168 | 147 | 130 | 148 | 151 |
| Germany | 293 | 279 | 311 | 391 | 639 | 808 | 3343 | 3301 | 3399 |
| Greenland | 40 | 66 | 437 | 288 | 866 | 533 | 1162 | 1129 | $747^{1, /}$ |
| Iceland | - | - | - | 19 | 82 | 7 | - | 1803 | 148 |
| Norway | 8 | 18 | 196 | 511 | 1120 | 1668 | 1881 | $1897{ }^{1}$ | 1253 |
| Russia | - | - | 5 | - | - | 10 | 424 | 37 | 52 |
| UK (Engl. and Wales) | 27 | 38 | 108 | 796 | 513 | 1405 | 264 | 218 | 190 |
| UK (Scotland) | - | - | 18 | 26 | 84 | 205 | 13 |  |  |
| United Kingdom | - | - | - | - | - | - | - |  |  |
| Total | 368 | 401 | 1075 | 2212 | 3472 | 4783 | 7218 | 8533 | 5940 |
| Working Group estimate | $736{ }^{2}$ | $875{ }^{\text { }}$ | $1176{ }^{4}$ | $2249{ }^{\text { }}$ | $3125{ }^{6}$ | 5077 | $7283{ }^{8}$ | 8558 |  |
|  |  |  |  |  |  |  |  |  |  |
| Country | 1999 | 2000 | $2001{ }^{1}$ | $2002{ }^{1}$ |  |  |  |  |  |
| Denmark |  |  |  |  |  |  |  |  |  |
| Faroe Islands | 2 |  |  |  |  |  |  |  |  |
| Germany | 3047 | 3243 | 2750 | 2117 |  |  |  |  |  |
| Greenland | $200^{1,4}$ | $1740{ }^{8}$ | $1553{ }^{9}$ |  |  |  |  |  |  |
| Iceland | 93 | 30 | 14280 |  |  |  |  |  |  |
| Ireland |  |  | 7 |  |  |  |  |  |  |
| Norway | 1100 | $1162^{\text {l }}$ | 1424 | 1280 |  |  |  |  |  |
| Portugal |  |  | 6 |  |  |  |  |  |  |
| Russia | 138 | 183 | 186 | 44 |  |  |  |  |  |
| Spain |  | 8 | 10 |  |  |  |  |  |  |
| UK (Engl. and Wales) | 226 | 262 | 100 |  |  |  |  |  |  |
| UK (Scotland) |  |  |  |  |  |  |  |  |  |
| United Kingdom |  |  |  | 202 |  |  |  |  |  |
| Total | 4806 | 6628 | 20316 | 3643 |  |  |  |  |  |
| Working Group estimate | $5376{ }^{11}$ | $6588{ }^{5}$ | $6588{ }^{6}$ | 6750 |  |  |  |  |  |

1) Provisional data
2)WG estimate includes additional catches as described in working Group reports for each year and in the report from 2001.
2) Includes $125 t$ by Faroe Islands and $206 t$ by Greenland.
3) Excluding 4732 t reported as area unknown.
4) Includes 1523 t by Norway, 102 t by Faroe Islands, 3343 t by Germany, 1910 t by Greenland, 180 t by Russia, as reported to Greenland authorities.
5) Includes 2849 t by Greenland, 142 t by Norway, 2750 t by Germany. Does not include 14280 t by Iceland as those are included in WG estimate of Va.
6) Excluding 138 t reported as area unknown.
7) Excluding 16 t reported as area unknown.
8) Excluding $20 t$ reported as area unknown

Table 6.1.5 GREENLAND HALIBUT. Nominal catches (tonnes) by countries in Sub-area XII, as officially reported to the ICES.

| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Faroe Islands |  | 47 |  |  |  |  |  |
| Norway | 2 |  |  |  |  |  |  |
| Total | 2 | 47 | - | - | - |  |  |
| WG estimate |  |  |  |  |  | 102 |  |

${ }^{1} 102 t$ by Faroe Islands as reported to Faroe Island authorities

Table 6.1.6. 2002 Catch statistics for Greenland halibut in V and XIV.
Working Group best estimates.

| Va | Long line | Trawl | Gill Net | Unknown | SUM | "Official" |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands |  |  |  |  | 0 |  |
| Germany, Fed. Rep. |  | 31 |  |  | 31 | 31 |
| Greenland |  |  |  | 424 | 424 | 424 |
| Iceland | 180 | 18307 | 772 |  | 19259 | 19233 |
| Norway |  |  |  |  | 0 |  |
| UK (E/W/NI) |  |  |  |  | 0 |  |
| UK (Scotland) |  |  |  |  | 0 |  |
| UK |  | 37 |  |  | 37 | 37 |
| Total | 180 | 18338 | 772 | 424 | 19714 | 19725 |
| Vb | Long line | Trawl | Gill Net | Unknown | SUM | "Official" |
| Faroe Islands |  | 449 | 1975 |  | 2424 |  |
| France |  |  |  | 20 | 20 | 20 |
| Germany Fed. Rep. |  |  |  |  | 0 |  |
| Norway |  |  |  | 48 | 48 | 48 |
| UK (England \& Wales) |  |  |  |  | 0 |  |
| UK (Scotland) |  |  |  |  | 0 |  |
| United Kingdom |  |  |  | 202 | 202 | 202 |
| Total | 0 | 449 | 1975 | 270 | 2694 | 270 |
| XII | Long line | Trawl | Gill Net | Unknown | SUM | SUM |
| Faroe Islands | 0 |  |  |  | 0 |  |
| Total | 0 | 0 | 0 | 102 | 102 | 0 |
| XIV | Long line | Trawl | Gill Net | Unknown | SUM | "Official" |
| Faroe Islands |  | 193 |  |  | 193 |  |
| EU (GER) |  | 2158 |  |  | 2158 | 2117 |
| Greenland |  | 2091 | 91 | 618 | 2800 |  |
| Iceland (outside 200 EEZ) |  |  |  |  | 0 |  |
| Norway (inside 200 EEZ) | 704 | 715 |  |  | 1419 | 1280 |
| Norway (outside 200 EEZ) |  |  |  |  | 0 |  |
| Portugal |  | 130 |  |  | 130 |  |
| Russia |  | 50 |  |  | 50 | 44 |
| Ireland |  |  |  |  | 0 |  |
| UK (England \& Wales) |  |  |  |  | 0 |  |
| UK (Scotland) |  |  |  |  | 0 |  |
| United Kingdom |  |  |  |  | 0 | 201 |
| Total | 704 | 5337 | 91 | 618 | 6750 | 3642 |


| Summary of catch by gear | Long line | Trawl | Gill Net | Unknown | SUM |
| :--- | ---: | ---: | ---: | ---: | ---: |
|  | 884 | 24124 | 2838 | 1414 | $\mathbf{2 9 2 6 0}$ |


| SUM |
| ---: |
| 23637 |

Table 6.2.1. CPUE indices of trawl fleets in Div. $\mathrm{Va}, \mathrm{Vb}$ and XIVb as derived from GLM multiplicative models.

| area | year | cpue | change in CPUE between years | landings | effort | \% change in effort between years |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Iceland Va | 1985 | 1.00 |  | 29,197 | 29 |  |
|  | 1986 | 0.91 | -9 | 31,027 | 34 | 7 |
|  | 1987 | 0.88 | -4 | 44,659 | 51 | 50 |
|  | 1988 | 0.96 | 9 | 49,379 | 51 | -6 |
|  | 1989 | 0.91 | -5 | 59,272 | 65 | 21 |
|  | 1990 | 0.74 | -19 | 37,308 | 50 | -10 |
|  | 1991 | 0.75 | 2 | 35,413 | 47 | -11 |
|  | 1992 | 0.65 | -14 | 31,978 | 49 | 19 |
|  | 1993 | 0.50 | -23 | 34,134 | 68 | 36 |
|  | 1994 | 0.41 | -19 | 28,608 | 70 | 18 |
|  | 1995 | 0.31 | -24 | 27,391 | 88 | 28 |
|  | 1996 | 0.26 | -17 | 22,073 | 85 | 14 |
|  | 1997 | 0.29 | 11 | 16,792 | 58 | -18 |
|  | 1998 | 0.46 | 61 | 10,595 | 23 | -56 |
|  | 1999 | 0.53 | 15 | 11,138 | 21 | -12 |
|  | 2000 | 0.59 | 10 | 14,519 | 25 | 22 |
|  | 2001 | 0.61 | 4 | 16,752 | 27 | -13 |
|  | 2002 | 0.47 | -24 | 19,714 | 42 | 54 |
| Greenland, XIVb | 1991 | 1.00 |  | 875 | 0.9 |  |
|  | 1992 | 0.87 | -13 | 1,176 | 1.4 | 54 |
|  | 1993 | 0.92 | 6 | 2,249 | 2.4 | 80 |
|  | 1994 | 1.18 | 28 | 3,125 | 2.6 | 9 |
|  | 1995 | 1.13 | -5 | 5,077 | 4.5 | 70 |
|  | 1996 | 1.14 | 2 | 7,283 | 6.4 | 41 |
|  | 1997 | 1.19 | 4 | 8,558 | 7.2 | 13 |
|  | 1998 | 1.20 | 0 | 5,940 | 5.0 | -31 |
|  | 1999 | 1.08 | -9 | 5,376 | 5.0 | 0 |
|  | 2000 | 1.18 | 8 | 6,588 | 5.6 | 13 |
|  | 2001 | 1.11 | -5 | 6,588 | 5.9 | 6 |
|  | 2002 | 1.11 | 0 | 6,750 | 6.1 | 3 |
| Faroe Islands, Vb | 1991 | 1.00 |  | 1,662 | 1.7 | 0 |
|  | 1992 | 1.02 | 2 | 2,269 | 2.2 | 34 |
|  | 1993 | 0.95 | -6 | 4,434 | 4.7 | 108 |
|  | 1994 | 0.90 | -6 | 5,225 | 5.8 | 25 |
|  | 1995 | 0.89 | -1 | 3,832 | 4.3 | -26 |
|  | 1996 | 0.88 | 0 | 6,469 | 7.3 | 70 |
|  | 1997 | 0.86 | -2 | 4,870 | 5.7 | -23 |
|  | 1998 | 0.88 | 2 | 3,825 | 4.3 | -23 |
|  | 1999 | 0.87 | -1 | 2,694 | 3.1 | -29 |
|  | 2000 | 0.87 | 0 | 5,092 | 5.9 | 89 |
|  | 2001 | 0.88 | 1 | 3,951 | 4.5 | -23 |
|  | 2002 | 0.89 | 1 | 2,694 | 3.0 | -33 |

Table 6.3.1

## GREENLAND HALIBUT ICES V+XIV

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|  | $\begin{aligned} & \text { Table } 1 \\ & \text { YEAR, } \end{aligned}$ | Catch numbers-at-age |  |  | Numbers*10**-3 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1975, | 1976, | 1977, | 1978, | 1979, | 1980, | 1981, | 1982, |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 5, | 120, | 43, | 0, | 23, | 29, | 47, | 26, | 8, |  |  |
|  | 6 , | 800, | 296, | 34, | 91, | 197, | 502, | 158, | 300, |  |  |
|  | 7, | 1775, | 584, | 671, | 347, | 1605, | 1536, | 580, | 1140, |  |  |
|  | 8, | 1782, | 621, | 1727, | 1037, | 2253, | 2630, | 1160, | 2451, |  |  |
|  | 9, | 1259, | 431, | 2289, | 1214, | 3090, | 3126, | 1430, | 2646, |  |  |
|  | 10, | 926, | 240, | 834, | 848, | 1693, | 2324, | 1764, | 2456, |  |  |
|  | 11, | 464, | 121, | 420, | 567 , | 880, | 1739, | 1299, | 1803, |  |  |
|  | 12, | 459, | 86, | 423, | 312, | 394, | 849, | 664, | 963, |  |  |
|  | 13, | 279, | 37, | 174, | 232, | 246, | 578, | 435, | 609, |  |  |
|  | 14, | 193, | 32, | 120, | 218, | 189, | 306, | 252, | 331, |  |  |
|  | 15, | 137, | 14, | 28, | 114, | 147, | 143, | 176, | 195, |  |  |
|  | 16, | 39, | 6, | 86, | 112, | 101, | 82, | 114, | 82, |  |  |
| 0 | TOTALNUM, | 8233, | 2511, | 6806, | 5115, | 10824, | 13862, | 8058, | 12984, |  |  |
|  | TONSLAND, | 23494, | 6045 , | 16578, | 14349, | 23616, | 31252, | 19239, | 32441, |  |  |
|  | SOPCOF \%, | 128, | 101, | 102, | 105, | 102, | 100, | 102, | 101, |  |  |
|  | Table 1 | Catch numbers-at-age |  |  | Numbers*10**-3 |  |  |  |  |  |  |
|  | YEAR, | 1983, | 1984, | 1985, | 1986, | 1987, | 1988, | 1989, | 1990, | 1991, | 1992, |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 5, | 10, | 83, | 125, | 245, | 182, | 129, | 499, | 188, | 289, | 17, |
|  | 6 , | 240, | 277, | 441, | 612, | 3123, | 742, | 1657, | 463, | 1225, | 421, |
|  | 7, | 1611, | 891, | 1018, | 1033, | 4863, | 2068, | 4485, | 1513, | 1797, | 2023, |
|  | 8, | 2651, | 2139, | 2295, | 1942, | 2586, | 2985, | 5961, | 3515, | 2866, | 3262, |
|  | 9, | 3060 , | 3568, | 3454, | 2983, | 2156, | 3166, | 5763, | 4186, | 2935, | 2646, |
|  | 10, | 2443, | 2800, | 2749, | 3097 , | 3476, | 2966, | 3246, | 3143, | 2074, | 3019, |
|  | 11, | 1693, | 1825, | 1452, | 1683, | 1847, | 1848, | 1601, | 1224, | 1130, | 1962, |
|  | 12, | 978, | 1134, | 627, | 820, | 1829, | 1761, | 1458, | 959, | 1072, | 1278, |
|  | 13, | 424, | 588, | 423, | 550, | 886, | 1851, | 1237, | 568, | 924, | 509, |
|  | 14, | 174, | 363, | 137, | 202, | 243, | 701, | 506, | 358, | 554, | 144, |
|  | 15, | 37, | 92, | 36, | 59, | 31, | 216, | 362, | 137, | 342, | 36, |
|  | 16, | 17, | 13, | 46, | 30, | 1, | 246, | 145, | 61, | 82, | 56, |
| 0 | TOTALNUM, | 13338, | 13773, | 12803, | 13256, | 21223, | 18679, | 26920, | 16315, | 15290, | 15373, |
|  | TONSLAND, | 30891, | 34024, | 32075, | 32984, | 46622, | 51118, | 61396, | 39326, | 37950, | 35423, |
|  | SOPCOF \%, | 102, | 100, | 103, | 101, | 98, | 101, | 100, | 100, | 101, | 100, |
|  | Table 1 | Catch numbers-at-age |  |  | Numbers*10**-3 |  |  |  |  |  |  |
|  | YEAR, | 1993, | 1994, | 1995, | 1996, | 1997, | 1998, | 1999, | 2000, | 2001, | 2002, |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 5, | 44, | 78, | 503, | 178, | 86, | 90, | 82, | 53, | 46, | 74, |
|  | 6, | 397, | 672, | 1587, | 1488, | 549, | 550, | 366, | 313, | 176, | 391, |
|  | 7, | 1896, | 2197, | 3031, | 2908, | 2723, | 1882, | 1363, | 1559, | 1016, | 2069, |
|  | 8, | 5024, | 3815, | 3287, | 3181, | 2579, | 2051, | 1606, | 2279, | 2975, | 3388, |
|  | 9, | 4324, | 3648, | 2608, | 2119, | 2331, | 1657, | 1828, | 1500, | 2387, | 2621, |
|  | 10, | 2859, | 2330, | 1963, | 1755, | 1247, | 1067, | 1287, | 1269, | 1574, | 1304, |
|  | 11, | 1539, | 1715, | 1548, | 1610, | 975, | 737, | 1018, | 1247, | 1550, | 897, |
|  | 12, | 1412, | 990, | 1132, | 1216, | 937, | 710, | 762, | 1016, | 1061, | 893, |
|  | 13, | 576, | 422, | 657, | 665, | 652, | 359, | 492, | 785, | 808, | 1047, |
|  | 14, | 136, | 371, | 444, | 548, | 374, | 195, | 231, | 786, | 379, | 719, |
|  | 15, | 135, | 168, | 240, | 238, | 282, | 150, | 137, | 545, | 422, | 309, |
|  | 16, | 7, | 177, | 211, | 323, | 262, | 106, | 119, | 176, | 123, | 469, |
| 0 | TOTALNUM, | 18349, | 16583, | 17211, | 16229, | 12997, | 9554, | 9291, | 11528, | 12517, | 14181, |
|  | TONSLAND, | 40817, | 36958, | 36300, | 35825, | 30267, | 20360, | 20371, | 26839, | 28284, | 28888, |
|  | SOPCOF \%, | 100, | 100, | 100, | 103, | 110, | 107, | 111, | 107, | 103, | 103, |

## Table 6.4.1 Catch and stock weight-at-age

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|  | Table 2 | Catch weights-at-age (kg) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR, | 1975, | 1976, | 1977, | 1978, | 1979, | 1980, | 1981, | 1982, |
|  | AGE |  |  |  |  |  |  |  |  |
|  | 5, | . 9680, | 1.1570, | 1.1570, | . 9680 , | . 9110, | 1.1250, | 1.0710, | 1.0100, |
|  | 6, | 1.1990, | 1.5850, | 1.0460, | 1.1990, | . 9420, | 1.2830, | 1.2570, | 1.3680, |
|  | 7, | 1.4230, | 1.7680, | 1.4290, | 1.4230, | 1.2780, | 1.4870, | 1.4400, | 1.6180, |
|  | 8, | 1.8540, | 2.1800, | 1.7940, | 1.8540, | 1.6760, | 1.7560, | 1.6600, | 1.9050, |
|  | 9, | 2.2560, | 2.5700, | 2.2280, | 2.2560, | 2.0720, | 2.1530, | 1.9670, | 2.1870, |
|  | 10, | 2.6070, | 3.0180 , | 2.6870, | 2.6070, | 2.3330, | 2.2790, | 2.2580, | 2.5160, |
|  | 11, | 3.0810, | 3.7300, | 3.0170, | 3.0810 , | 2.7230, | 2.4980, | 2.5150, | 2.7610, |
|  | 12, | 3.5910, | 4.0520, | 3.9140, | 3.5910, | 3.2970, | 3.0590, | 2.9500, | 3.1290, |
|  | 13, | 4.6040, | 4.8150, | 4.0400, | 4.6040, | 3.9850, | 3.7830, | 3.4500, | 3.7850, |
|  | 14, | 4.6950, | 5.3480, | 4.7140, | 4.6950, | 4.6680, | 4.5070, | 4.0330, | 4.4750, |
|  | 15, | 5.1510, | 5.7520, | 5.4010, | 5.1510, | 4.7920, | 5.1390, | 4.6520, | 4.9850, |
|  | 16, | 5.8930, | 6.2270 , | 5.0540, | 5.8930, | 5.2290, | 5.6330, | 4.7140, | 5.6100, |
|  | SOPCOFAC, | 1.2794, | 1.0068, | 1.0227, | 1.0471, | 1.0187, | . 9975 , | 1.0189, | 1.0104, |


|  | Table 2 | Catch weights-at-age (kg) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR, | 1983, | 1984, | 1985, | 1986, | 1987, | 1988, | 1989, | 1990, | 1991, | 1992, |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 5, | . 9840 , | . 9420 , | . 9950, | 1.0300, | 1.0300, | 1.1290, | . 8420 , | 1.0290, | 1.0010, | 1.0160, |
|  | 6 , | 1.3380, | 1.2750, | 1.2300, | 1.2380, | 1.2180, | 1.3040, | 1.0470, | 1.2100, | 1.2470, | 1.2560, |
|  | 7, | 1.5770, | 1.5920, | 1.6300, | 1.4990, | 1.5330, | 1.5410, | 1.4250, | 1.5720, | 1.4720, | 1.4010, |
|  | 8, | 1.8480, | 1.8170, | 1.9510, | 1.9370, | 1.8240, | 1.7700, | 1.7270, | 1.7900, | 1.8100, | 1.7180, |
|  | 9, | 2.1590, | 2.2400, | 2.3670, | 2.3630, | 2.1870, | 2.2360, | 2.1250, | 2.1260, | 2.0880, | 2.0490 , |
|  | 10, | 2.4340, | 2.4610, | 2.6370, | 2.6310 , | 2.6660, | 2.6830, | 2.6370, | 2.5360, | 2.4400, | 2.4360, |
|  | 11, | 2.6030, | 2.8350, | 2.8290, | 2.8480, | 2.9960, | 3.0820, | 3.2200, | 3.2140, | 2.9350, | 2.8680 , |
|  | 12, | 3.0340, | 3.2620, | 3.3530, | 3.3350, | 3.5950, | 3.6240, | 3.7330, | 3.6930, | 3.7370, | 3.4780 , |
|  | 13, | 3.7840, | 3.9620, | 4.0060, | 4.0390, | 4.4310, | 4.3120, | 4.1350, | 4.4480, | 4.4010, | 4.5100, |
|  | 14, | 4.4460, | 4.9360, | 4.7920, | 4.9250, | 5.1400, | 5.0980 , | 5.3800, | 5.1970, | 5.0220, | 4.6810, |
|  | 15, | 4.7510, | 5.2300, | 5.2310, | 5.4660, | 5.7640, | 5.2130, | 6.5690 , | 5.8910, | 5.9910, | 6.0100 , |
|  | 16, | 6.2090, | 6.9680, | 6.3230, | 5.7640, | 5.7640, | 5.7640, | 6.4970, | 6.0490, | 6.4120, | 5.1280 , |
| 0 | SOPCOFAC, | 1.0176, | . 9953 , | 1.0258, | 1.0069, | . 9792 , | 1.0063, | . 9999 , | . 9998 , | 1.0097, | 1.0033 , |


|  | Table 2 | Catch weights-at-age (kg) |  |  |  | 1997, | 1998, | 1999, | 2000, | 2001, | 2002, |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR, | 1993, | 1994, | 1995, | 1996, |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 5, | . 9910 , | 1.1630, | . 9500 , | 1.1010, | . 9190, | . 8070 , | . 8610 , | . 7700 , | . 7320 , | . 7210 , |
|  | 6, | 1.2490, | 1.2540, | 1.2130, | 1.1240, | 1.1070, | 1.0860, | . 9530 , | . 9680 , | . 8650 , | . 9850 , |
|  | 7, | 1.4010, | 1.4880, | 1.4130, | 1.3460, | 1.3340, | 1.3630, | 1.2880, | 1.2720, | 1.2190, | 1.3550, |
|  | 8, | 1.6850, | 1.7360, | 1.7030, | 1.6490, | 1.6400, | 1.6580, | 1.5650, | 1.6070, | 1.5770, | 1.6290, |
|  | 9, | 1.9820, | 2.1500, | 2.0280, | 1.9250, | 1.8810, | 1.8860, | 1.7390, | 1.7690, | 1.8510, | 1.9910, |
|  | 10, | 2.4250, | 2.3520, | 2.2790, | 2.3420, | 2.2400, | 2.1670, | 2.0120, | 2.1220, | 2.2020, | 2.0890, |
|  | 11, | 2.9520, | 2.7360, | 2.6430, | 2.5950, | 2.5380, | 2.4150, | 2.3510, | 2.3140, | 2.3990, | 2.3270, |
|  | 12, | 3.4290, | 3.0820, | 2.9920, | 3.0130, | 2.8460, | 2.8440, | 2.6340, | 2.7220, | 2.7000, | 2.4720, |
|  | 13, | 4.4790, | 3.6070, | 3.5680, | 3.5150, | 3.3850, | 3.1730, | 3.0310, | 3.0100, | 3.3110, | 2.7220, |
|  | 14, | 6.0430 , | 4.2420, | 4.0680, | 4.1230, | 4.3590, | 4.2370, | 3.5320, | 3.4230, | 4.0720, | 2.8710, |
|  | 15, | 5.8320, | 5.2930, | 5.3020, | 4.9960, | 4.8510, | 4.6560, | 3.8740, | 4.0660, | 4.5500, | 2.8660, |
|  | 16, | 5.5120, | 6.0870, | 5.6860, | 5.6930, | 5.0910, | 5.0800, | 4.9370, | 4.5730, | 5.8670, | 2.9010, |
| 0 | SOPCOFAC, | 1.0010, | 1.0001, | 1.0042, | 1.0329, | 1.1044, | 1.0674, | 1.1142, | 1.0710, | 1.0306, | 1.0259, |

Table 6.7.2.1
Output from ASPIC model on CPUE series in Div. Va, total catches V+XIV and Icelandic fall survey indices.

## Greenland halibut XIV and V

Page 1

ASPIC -- A Surplus-Production Model Including Covariates (BETA Ver. 4.45)
Author: Michael H. Prager; NOAA Center for Coastal Fisheries and Habitat Research 101 Pivers Island Road; Beaufort, North Carolina 28516 USA Mike.Prager@noaa.gov

FIT program mode
LOGISTIC model mode
YLD conditioning
SSE optimization
Reference: Prager, M. H. 1994. A suite of extensions to a nonequilibrium
ASPIC User's Manual is available surplus-production model. Fishery Bulletin 92: 374-389.
gratis from the author.
CONTROL PARAMETERS USED (FROM INPUT FILE) Input file: ghl8502-new.inp
Operation of ASPIC: Fit logistic model by direct optimization.
Number of years analyzed: $18 \quad$ Number of bootstrap trials:
Number of years analyzed
Lower bound on MSY:
$5.000 \mathrm{E}+03$
Number of data series:
Upper bound on MSY:
$1.000 \mathrm{E}+09$
Objective function: Least squares
Upper bound on MSY: . $000 \mathrm{E}+09$
Relative conv. criterion (simplex):
Relative conv. criterion (restart):
east squares

Relative conv. criterion (effort): $1.000 \mathrm{E}-04$
$\begin{array}{ll}\text { Maximum } F \text { allowed in fitting: } & 6.000 \\ \text { Identical convergences required in fitting: } & 5\end{array}$
Upper bound on K :
Random number seed:

PROGRAM STATUS INFORMATION (NON-BOOTSTRAPPED ANALYSIS)
Monte Carlo search mode, trials:

Normal convergence.
Number of restarts required for convergence: 34

CORRELATION AMONG INPUT SERIES EXPRESSED AS CPUE (NUMBER OF PAIRWISE OBSERVATIONS BELOW)


GOODNESS-OF-FIT AND WEIGHTING FOR NON-BOOTSTRAPPED ANALYSIS


MANAGEMENT and DERIVED PARAMETER ESTIMATES (NON-BOOTSTRAPPED)

| Parameter |  | Estimate | Logistic formula | General formula |
| :---: | :---: | :---: | :---: | :---: |
| MSY | Maximum sustainable yield | $3.481 \mathrm{E}+04$ |  |  |
| $\mathrm{B}_{\text {MSY }}$ | Stock biomass giving MSY | $1.138 \mathrm{E}+05$ | K/2 | K * n ** (1/(1-n) ) |
| $\mathbf{F}_{\text {MSY }}$ | Fishing mortality rate at MSY | $3.059 \mathrm{E}-01$ | $\mathrm{MSY} / \mathrm{B}_{\text {MSY }}$ | $\mathrm{MSY} / \mathrm{B}_{\text {MSY }}$ |
| n | Exponent in production function | $2.000 \mathrm{E}+00$ | ---- |  |
| g | Fletcher's gamma | $4.000 \mathrm{E}+00$ | ---- | [ n **(n/(n-1))]/[n-1] |
| B. $/ \mathrm{B}_{\text {MSY }}$ | Ratio: B (2003) / $\mathrm{B}_{\text {msY }}$ | $7.761 \mathrm{E}-01$ | ---- | ---- |
| F. $/ \mathrm{F}_{\text {MSY }}$ | Ratio: F (2002)/F ${ }_{\text {MSY }}$ | $1.105 \mathrm{E}+00$ |  |  |
| $\mathbf{F}_{\text {MSY }} / \mathrm{F}$. | Ratio: $\mathbf{F}_{\text {MSY }} / \mathrm{F}(2002)$ | $9.053 \mathrm{E}-01$ | ---- | ---- |
| Y. ( $\mathrm{F}_{\text {MSY }}$ ) | Yield available at $\mathbf{F}_{\text {MSY }}$ in 2003 ...as proportion of MSY | $\begin{aligned} & 2.702 \mathrm{E}+04 \\ & 7.761 \mathrm{E}-01 \end{aligned}$ | MSY ${ }^{\text {* }}$. $/ \mathrm{B}_{\text {MSY }}$ | MSY*B. $\mathrm{B}_{\text {MSY }}$ |
| Ye. | Equilibrium yield available in 2003 ...as proportion of MSY | $\begin{aligned} & 3.307 \mathrm{E}+04 \\ & 9.499 \mathrm{E}-01 \end{aligned}$ | $4 * M S Y *(B / K-(B / K) * * 2)$ | g MSY* (B/K-(B/K)**n) |
| $\mathrm{F}_{\mathrm{MSY}}$ (1) | input CPUE indices | $4.905 \mathrm{E}+01$ | $\boldsymbol{F}_{\text {MSY }} / \mathrm{q}(1)$ | $\boldsymbol{F}_{\text {MSY }} / \mathrm{q}(1)$ |

Table 6.7.2.1 (Cont'd)
ESTIMATED POPULATION TRAJECTORY (NON-BOOTSTRAPPED)

|  | Year | Estimated total | Estimated starting | Estimated average | Observed total | Model total | Estimated surplus | Ratio of F mort | Ratio of biomass |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Obs | or ID | F mort | biomass | biomass | yield | yield | production | to $\mathbf{F}_{\text {MSY }}$ | to $\mathbf{B}_{\text {MSY }}$ |
| 1 | 1985 | 0.190 | 1.712E+05 | 1. $684 \mathrm{E}+05$ | $3.208 \mathrm{E}+04$ | $3.208 \mathrm{E}+04$ | $2.680 \mathrm{E}+04$ | $6.228 \mathrm{E}-01$ | $1.504 \mathrm{E}+00$ |
| 2 | 1986 | 0.202 | $1.660 \mathrm{E}+05$ | 1. $634 \mathrm{E}+05$ | $3.298 \mathrm{E}+04$ | $3.298 \mathrm{E}+04$ | $2.820 \mathrm{E}+04$ | $6.601 \mathrm{E}-01$ | 1.458E+00 |
| 3 | 1987 | 0.306 | 1.612E+05 | $1.526 \mathrm{E}+05$ | $4.662 \mathrm{E}+04$ | $4.662 \mathrm{E}+04$ | $3.072 \mathrm{E}+04$ | 9.992E-01 | $1.416 \mathrm{E}+00$ |
| 4 | 1988 | 0.377 | $1.453 \mathrm{E}+05$ | 1.357E+05 | $5.112 \mathrm{E}+04$ | $5.112 \mathrm{E}+04$ | $3.345 \mathrm{E}+04$ | $1.231 \mathrm{E}+00$ | $1.276 \mathrm{E}+00$ |
| 5 | 1989 | 0.543 | $1.276 \mathrm{E}+05$ | $1.130 \mathrm{E}+05$ | $6.140 \mathrm{E}+04$ | $6.140 \mathrm{E}+04$ | $3.465 \mathrm{E}+04$ | $1.776 \mathrm{E}+00$ | 1.121E+00 |
| 6 | 1990 | 0.401 | $1.009 \mathrm{E}+05$ | $9.814 \mathrm{E}+04$ | $3.933 \mathrm{E}+04$ | $3.933 \mathrm{E}+04$ | $3.415 \mathrm{E}+04$ | 1.310E+00 | $8.863 \mathrm{E}-01$ |
| 7 | 1991 | 0.406 | $9.569 \mathrm{E}+04$ | $9.346 \mathrm{E}+04$ | $3.795 \mathrm{E}+04$ | $3.795 \mathrm{E}+04$ | $3.369 \mathrm{E}+04$ | 1.328E+00 | $8.407 \mathrm{E}-01$ |
| 8 | 1992 | 0.392 | $9.143 \mathrm{E}+04$ | $9.034 \mathrm{E}+04$ | $3.542 \mathrm{E}+04$ | $3.542 \mathrm{E}+04$ | $3.333 \mathrm{E}+04$ | 1.282E+00 | $8.033 \mathrm{E}-01$ |
| 9 | 1993 | 0.480 | $8.934 \mathrm{E}+04$ | $8.498 \mathrm{E}+04$ | $4.082 \mathrm{E}+04$ | $4.082 \mathrm{E}+04$ | $3.256 \mathrm{E}+04$ | $1.570 \mathrm{E}+00$ | $7.849 \mathrm{E}-01$ |
| 10 | 1994 | 0.473 | $8.108 \mathrm{E}+04$ | $7.817 \mathrm{E}+04$ | $3.696 \mathrm{E}+04$ | $3.696 \mathrm{E}+04$ | $3.139 \mathrm{E}+04$ | $1.546 \mathrm{E}+00$ | $7.124 \mathrm{E}-01$ |
| 11 | 1995 | 0.502 | $7.551 \mathrm{E}+04$ | $7.231 \mathrm{E}+04$ | $3.630 \mathrm{E}+04$ | $3.630 \mathrm{E}+04$ | $3.017 \mathrm{E}+04$ | $1.641 \mathrm{E}+00$ | $6.635 \mathrm{E}-01$ |
| 12 | 1996 | 0.546 | $6.939 \mathrm{E}+04$ | $6.557 \mathrm{E}+04$ | $3.582 \mathrm{E}+04$ | $3.582 \mathrm{E}+04$ | $2.854 \mathrm{E}+04$ | $1.786 \mathrm{E}+00$ | $6.096 \mathrm{E}-01$ |
| 13 | 1997 | 0.500 | $6.210 \mathrm{E}+04$ | $6.050 \mathrm{E}+04$ | $3.027 \mathrm{E}+04$ | $3.027 \mathrm{E}+04$ | $2.717 \mathrm{E}+04$ | $1.636 \mathrm{E}+00$ | $5.457 \mathrm{E}-01$ |
| 14 | 1998 | 0.324 | $5.901 \mathrm{E}+04$ | $6.276 \mathrm{E}+04$ | $2.036 \mathrm{E}+04$ | $2.036 \mathrm{E}+04$ | $2.779 \mathrm{E}+04$ | $1.061 \mathrm{E}+00$ | $5.184 \mathrm{E}-01$ |
| 15 | 1999 | 0.286 | $6.644 \mathrm{E}+04$ | $7.126 \mathrm{E}+04$ | $2.037 \mathrm{E}+04$ | $2.037 \mathrm{E}+04$ | 2.992E+04 | $9.346 \mathrm{E}-01$ | $5.838 \mathrm{E}-01$ |
| 16 | 2000 | 0.343 | $7.599 \mathrm{E}+04$ | $7.835 \mathrm{E}+04$ | $2.684 \mathrm{E}+04$ | $2.684 \mathrm{E}+04$ | $3.143 \mathrm{E}+04$ | 1.120E+00 | $6.677 \mathrm{E}-01$ |
| 17 | 2001 | 0.339 | $8.058 \mathrm{E}+04$ | $8.274 \mathrm{E}+04$ | $2.802 \mathrm{E}+04$ | $2.802 \mathrm{E}+04$ | $3.221 \mathrm{E}+04$ | $1.107 \mathrm{E}+00$ | $7.080 \mathrm{E}-01$ |
| 18 | 2002 | 0.338 | $8.477 \mathrm{E}+04$ | 8.661E+04 | $2.926 \mathrm{E}+04$ | $2.926 \mathrm{E}+04$ | $3.282 \mathrm{E}+04$ | $1.105 \mathrm{E}+00$ | $7.448 \mathrm{E}-01$ |
| 19 | 2003 |  | $8.833 \mathrm{E}+04$ |  |  |  |  |  | $7.761 \mathrm{E}-01$ |

RESULTS FOR DATA SERIES \# 1 (NON-BOOTSTRAPPED)

|  |  | Observed | Estimated | Estim | Observed | Model | Resid in |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Obs | Year | CPUE | CPUE | F | yield | yield | log scale |
| 1 | 1985 | $1.000 \mathrm{E}+03$ | $1.050 \mathrm{E}+03$ | 0.1905 | $3.208 \mathrm{E}+04$ | $3.208 \mathrm{E}+04$ | 0.04867 |
| 2 | 1986 | $9.130 \mathrm{E}+02$ | $1.019 \mathrm{E}+03$ | 0.2019 | $3.298 \mathrm{E}+04$ | $3.298 \mathrm{E}+04$ | 0.10954 |
| 3 | 1987 | $8.800 \mathrm{E}+02$ | $9.511 \mathrm{E}+02$ | 0.3056 | $4.662 \mathrm{E}+04$ | $4.662 \mathrm{E}+04$ | 0.07775 |
| 4 | 1988 | $9.610 \mathrm{E}+02$ | $8.462 \mathrm{E}+02$ | 0.3766 | $5.112 \mathrm{E}+04$ | $5.112 \mathrm{E}+04$ | -0.12718 |
| 5 | 1989 | $9.140 \mathrm{E}+02$ | $7.048 \mathrm{E}+02$ | 0.5431 | $6.140 \mathrm{E}+04$ | $6.140 \mathrm{E}+04$ | -0.25987 |
| 6 | 1990 | $7.420 \mathrm{E}+02$ | $6.119 \mathrm{E}+02$ | 0.4007 | $3.933 \mathrm{E}+04$ | $3.933 \mathrm{E}+04$ | -0.19275 |
| 7 | 1991 | $7.540 \mathrm{E}+02$ | $5.827 \mathrm{E}+02$ | 0.4061 | $3.795 \mathrm{E}+04$ | $3.795 \mathrm{E}+04$ | -0.25773 |
| 8 | 1992 | $6.520 \mathrm{E}+02$ | $5.633 \mathrm{E}+02$ | 0.3921 | $3.542 \mathrm{E}+04$ | $3.542 \mathrm{E}+04$ | -0.14630 |
| 9 | 1993 | $5.040 \mathrm{E}+02$ | $5.299 \mathrm{E}+02$ | 0.4803 | $4.082 \mathrm{E}+04$ | $4.082 \mathrm{E}+04$ | 0.05007 |
| 10 | 1994 | $4.100 \mathrm{E}+02$ | $4.874 \mathrm{E}+02$ | 0.4728 | $3.696 \mathrm{E}+04$ | $3.696 \mathrm{E}+04$ | 0.17286 |
| 11 | 1995 | $3.120 \mathrm{E}+02$ | $4.508 \mathrm{E}+02$ | 0.5020 | $3.630 \mathrm{E}+04$ | $3.630 \mathrm{E}+04$ | 0.36808 |
| 12 | 1996 | $2.590 \mathrm{E}+02$ | $4.088 \mathrm{E}+02$ | 0.5464 | $3.582 \mathrm{E}+04$ | $3.582 \mathrm{E}+04$ | 0.45647 |
| 13 | 1997 | $2.880 \mathrm{E}+02$ | $3.772 \mathrm{E}+02$ | 0.5003 | $3.027 \mathrm{E}+04$ | $3.027 \mathrm{E}+04$ | 0.26986 |
| 14 | 1998 | $4.640 \mathrm{E}+02$ | $3.913 \mathrm{E}+02$ | 0.3244 | $2.036 \mathrm{E}+04$ | $2.036 \mathrm{E}+04$ | -0.17047 |
| 15 | 1999 | $5.340 \mathrm{E}+02$ | $4.443 \mathrm{E}+02$ | 0.2859 | $2.037 \mathrm{E}+04$ | $2.037 \mathrm{E}+04$ | -0.18385 |
| 16 | 2000 | $5.890 \mathrm{E}+02$ | $4.885 \mathrm{E}+02$ | 0.3426 | $2.684 \mathrm{E}+04$ | $2.684 \mathrm{E}+04$ | -0.18712 |
| 17 | 2001 | $6.120 \mathrm{E}+02$ | $5.159 \mathrm{E}+02$ | 0.3387 | $2.802 \mathrm{E}+04$ | $2.802 \mathrm{E}+04$ | -0.17090 |
| 18 | 2002 | $4.680 \mathrm{E}+02$ | $5.400 \mathrm{E}+02$ | 0.3379 | $2.926 \mathrm{E}+04$ | $2.926 \mathrm{E}+04$ | 0.14307 |



Table 6.7.2.1 (Cont'd)
RESULTS FOR DATA SERIES \# 2 (NON-BOOTSTRAPPED)

| Data type I2: End-of-year biomass index |  |  |  |  |  |  |  | Series weight: | 1.000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Obs | Year | Observed effort | Estimated effort | Estim | Observed index | Model index | Resid in log index |  |  |
| 1 | 1985 | $0.000 \mathrm{E}+00$ | $0.000 \mathrm{E}+00$ | 0.0 | * | $1.004 \mathrm{E}+03$ | 0.00000 |  |  |
| 2 | 1986 | $0.000 \mathrm{E}+00$ | $0.000 \mathrm{E}+00$ | 0.0 | * | $9.749 \mathrm{E}+02$ | 0.00000 |  |  |
| 3 | 1987 | $0.000 \mathrm{E}+00$ | $0.000 \mathrm{E}+00$ | 0.0 | * | $8.788 \mathrm{E}+02$ | 0.00000 |  |  |
| 4 | 1988 | $0.000 \mathrm{E}+00$ | $0.000 \mathrm{E}+00$ | 0.0 | * | $7.719 \mathrm{E}+02$ | 0.00000 |  |  |
| 5 | 1989 | $0.000 \mathrm{E}+00$ | $0.000 \mathrm{E}+00$ | 0.0 | * | $6.101 \mathrm{E}+02$ | 0.00000 |  |  |
| 6 | 1990 | $0.000 \mathrm{E}+00$ | $0.000 \mathrm{E}+00$ | 0.0 | * | $5.788 \mathrm{E}+02$ | 0.00000 |  |  |
| 7 | 1991 | $0.000 \mathrm{E}+00$ | $0.000 \mathrm{E}+00$ | 0.0 | * | $5.531 \mathrm{E}+02$ | 0.00000 |  |  |
| 8 | 1992 | $0.000 \mathrm{E}+00$ | $0.000 \mathrm{E}+00$ | 0.0 | * | $5.404 \mathrm{E}+02$ | 0.00000 |  |  |
| 9 | 1993 | $0.000 \mathrm{E}+00$ | $0.000 \mathrm{E}+00$ | 0.0 | * | $4.905 \mathrm{E}+02$ | 0.00000 |  |  |
| 10 | 1994 | $0.000 \mathrm{E}+00$ | $0.000 \mathrm{E}+00$ | 0.0 | * | $4.568 \mathrm{E}+02$ | 0.00000 |  |  |
| 11 | 1995 | $0.000 \mathrm{E}+00$ | $0.000 \mathrm{E}+00$ | 0.0 | * | $4.197 \mathrm{E}+02$ | 0.00000 |  |  |
| 12 | 1996 | $1.000 \mathrm{E}+00$ | $1.000 \mathrm{E}+00$ | 0.0 | $3.460 \mathrm{E}+02$ | $3.757 \mathrm{E}+02$ | -0.08224 |  |  |
| 13 | 1997 | $1.000 \mathrm{E}+00$ | $1.000 \mathrm{E}+00$ | 0.0 | $4.140 \mathrm{E}+02$ | $3.569 \mathrm{E}+02$ | 0.14834 |  |  |
| 14 | 1998 | $1.000 \mathrm{E}+00$ | $1.000 \mathrm{E}+00$ | 0.0 | $4.200 \mathrm{E}+02$ | $4.019 \mathrm{E}+02$ | 0.04410 |  |  |
| 15 | 1999 | $1.000 \mathrm{E}+00$ | $1.000 \mathrm{E}+00$ | 0.0 | $5.280 \mathrm{E}+02$ | $4.597 \mathrm{E}+02$ | 0.13859 |  |  |
| 16 | 2000 | $1.000 \mathrm{E}+00$ | $1.000 \mathrm{E}+00$ | 0.0 | $3.960 \mathrm{E}+02$ | $4.874 \mathrm{E}+02$ | -0.20770 |  |  |
| 17 | 2001 | $1.000 \mathrm{E}+00$ | $1.000 \mathrm{E}+00$ | 0.0 | $5.570 \mathrm{E}+02$ | $5.128 \mathrm{E}+02$ | 0.08276 |  |  |
| 18 | 2002 | $1.000 \mathrm{E}+00$ | $1.000 \mathrm{E}+00$ | 0.0 | $4.720 \mathrm{E}+02$ | $5.343 \mathrm{E}+02$ | -0.12395 |  |  |

* Asterisk indicates missing value(s).


Table 6.7.3 Trajectories from ASPIC assuming 2003 catch eq 30 kt and $\mathrm{F}_{2004-2012}$ eq $\mathbf{F}_{\mathrm{pa}}\left(\sim 2 / 3 \mathbf{F}_{\mathrm{MSY}}\right)$.
Results from ASPICP.EXE, version 3.10

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Project $2 / 3 \mathbf{F}_{\mathrm{pa}}$

USER CONTROL INFORMATION (FROM INPUT FILE)

|  |  |  | aspic.bio |
| :---: | :---: | :---: | :---: |
| Name of biomass (BIO) file <br> Name of output file (this file) |  |  | ghlboot_catch23Fpa . out |
| Production-model type |  |  | Logistic |
| Number of years of projections |  |  | 10 |
| Type of confidence intervals |  |  | Bias-corrected percentile |
| Confidence interval smoothing |  |  | ON |
| Year | Input data | User data type |  |
| 2003 | $3.000 \mathrm{E}+04$ | TAC |  |
| 2004 | $6.000 \mathrm{E}-01$ | F/F (2002) |  |
| 2005 | $6.000 \mathrm{E}-01$ | F/F (2002) |  |
| 2006 | $6.000 \mathrm{E}-01$ | F/F(2002) |  |
| 2007 | $6.000 \mathrm{E}-01$ | F/F (2002) |  |
| 2008 | $6.000 \mathrm{E}-01$ | F/F (2002) |  |
| 2009 | $6.000 \mathrm{E}-01$ | F/F (2002) |  |
| 2010 | $6.000 \mathrm{E}-01$ | F/F (2002) |  |
| 2011 | $6.000 \mathrm{E}-01$ | F/F (2002) |  |
| 2012 | $6.000 \mathrm{E}-01$ | F/F(2002) |  |

TRAJECTORY OF RELATIVE BIOMASS B/B $\mathbf{B}_{\text {MSY }}$ (BOOTSTRAPPED)

| Year | $\begin{array}{r} \text { Point } \\ \text { estimate } \end{array}$ | Estimated bias | Relative bias | Approx 80\% lower CL | $\begin{aligned} & \text { Approx 80\% } \\ & \text { upper CL } \end{aligned}$ | Approx 50\% lower CL | Approx 50\% upper CL | Interquartile range | Relative IQ range |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | $1.503 \mathrm{E}+00$ | $1.669 \mathrm{E}-01$ | 11.10\% | $7.412 \mathrm{E}-01$ | $2.493 \mathrm{E}+00$ | $1.022 \mathrm{E}+00$ | $1.917 \mathrm{E}+00$ | 8.951E-01 | 0.595 |
| 1986 | $1.457 \mathrm{E}+00$ | $5.695 \mathrm{E}-02$ | 3.91\% | $8.132 \mathrm{E}-01$ | $2.040 \mathrm{E}+00$ | $1.076 \mathrm{E}+00$ | $1.713 \mathrm{E}+00$ | 6.379E-01 | 0.438 |
| 1987 | $1.416 \mathrm{E}+00$ | $1.692 \mathrm{E}-02$ | 1.20\% | $9.029 \mathrm{E}-01$ | $1.787 \mathrm{E}+00$ | $1.104 \mathrm{E}+00$ | $1.574 \mathrm{E}+00$ | $4.698 \mathrm{E}-01$ | 0.332 |
| 1988 | $1.277 \mathrm{E}+00$ | $1.815 \mathrm{E}-03$ | 0.14\% | $8.736 \mathrm{E}-01$ | $1.538 \mathrm{E}+00$ | $1.054 \mathrm{E}+00$ | $1.393 \mathrm{E}+00$ | 3.392E-01 | 0.266 |
| 1989 | $1.122 \mathrm{E}+00$ | -2.271E-03 | -0.20\% | $8.340 \mathrm{E}-01$ | $1.351 \mathrm{E}+00$ | $9.777 \mathrm{E}-01$ | $1.235 \mathrm{E}+00$ | $2.573 \mathrm{E}-01$ | 0.229 |
| 1990 | 8.889E-01 | -5.198E-03 | -0.58\% | $6.850 \mathrm{E}-01$ | $1.168 \mathrm{E}+00$ | $7.945 \mathrm{E}-01$ | $1.022 \mathrm{E}+00$ | $2.278 \mathrm{E}-01$ | 0.256 |
| 1991 | $8.436 \mathrm{E}-01$ | -4.350E-03 | -0.52\% | $6.558 \mathrm{E}-01$ | $1.100 \mathrm{E}+00$ | $7.573 \mathrm{E}-01$ | $9.632 \mathrm{E}-01$ | $2.058 \mathrm{E}-01$ | 0.244 |
| 1992 | $8.066 \mathrm{E}-01$ | -2.391E-03 | -0.30\% | $6.368 \mathrm{E}-01$ | $1.050 \mathrm{E}+00$ | $7.323 \mathrm{E}-01$ | $9.206 \mathrm{E}-01$ | $1.883 \mathrm{E}-01$ | 0.233 |
| 1993 | $7.879 \mathrm{E}-01$ | $1.374 \mathrm{E}-03$ | $0.17 \%$ | $6.292 \mathrm{E}-01$ | $1.006 \mathrm{E}+00$ | $7.196 \mathrm{E}-01$ | $8.836 \mathrm{E}-01$ | 1.640E-01 | 0.208 |
| 1994 | $7.120 \mathrm{E}-01$ | $4.606 \mathrm{E}-03$ | 0.65\% | $5.669 \mathrm{E}-01$ | $9.320 \mathrm{E}-01$ | $6.436 \mathrm{E}-01$ | $8.063 \mathrm{E}-01$ | $1.627 \mathrm{E}-01$ | 0.228 |
| 1995 | $6.609 \mathrm{E}-01$ | $8.335 \mathrm{E}-03$ | 1.26\% | $5.292 \mathrm{E}-01$ | $8.873 \mathrm{E}-01$ | $5.994 \mathrm{E}-01$ | $7.586 \mathrm{E}-01$ | $1.592 \mathrm{E}-01$ | 0.241 |
| 1996 | $6.067 \mathrm{E}-01$ | $1.176 \mathrm{E}-02$ | 1.94\% | $4.874 \mathrm{E}-01$ | $8.634 \mathrm{E}-01$ | $5.473 \mathrm{E}-01$ | $7.142 \mathrm{E}-01$ | $1.668 \mathrm{E}-01$ | 0.275 |
| 1997 | $5.418 \mathrm{E}-01$ | $1.311 \mathrm{E}-02$ | 2.42\% | $4.157 \mathrm{E}-01$ | $8.327 \mathrm{E}-01$ | $4.710 \mathrm{E}-01$ | $6.637 \mathrm{E}-01$ | $1.927 \mathrm{E}-01$ | 0.356 |
| 1998 | $5.185 \mathrm{E}-01$ | $1.184 \mathrm{E}-02$ | 2.28\% | $3.692 \mathrm{E}-01$ | $8.276 \mathrm{E}-01$ | $4.334 \mathrm{E}-01$ | $6.550 \mathrm{E}-01$ | $2.215 \mathrm{E}-01$ | 0.427 |
| 1999 | $5.830 \mathrm{E}-01$ | $1.051 \mathrm{E}-02$ | 1.80\% | $4.060 \mathrm{E}-01$ | $8.767 \mathrm{E}-01$ | $4.822 \mathrm{E}-01$ | $7.160 \mathrm{E}-01$ | $2.338 \mathrm{E}-01$ | 0.401 |
| 2000 | 6.659E-01 | $1.050 \mathrm{E}-02$ | 1.58\% | $4.687 \mathrm{E}-01$ | $9.457 \mathrm{E}-01$ | $5.577 \mathrm{E}-01$ | $8.110 \mathrm{E}-01$ | 2.533E-01 | 0.380 |
| 2001 | $7.077 \mathrm{E}-01$ | $1.090 \mathrm{E}-02$ | 1.54\% | $4.927 \mathrm{E}-01$ | $9.769 \mathrm{E}-01$ | $5.891 \mathrm{E}-01$ | $8.349 \mathrm{E}-01$ | 2.459E-01 | 0.347 |
| 2002 | $7.439 \mathrm{E}-01$ | $1.193 \mathrm{E}-02$ | 1.60\% | $4.873 \mathrm{E}-01$ | $9.922 \mathrm{E}-01$ | $6.005 \mathrm{E}-01$ | $8.653 \mathrm{E}-01$ | $2.649 \mathrm{E}-01$ | 0.356 |
| 2003 | $7.797 \mathrm{E}-01$ | $1.376 \mathrm{E}-02$ | 1.76\% | $4.991 \mathrm{E}-01$ | $1.015 \mathrm{E}+00$ | $6.252 \mathrm{E}-01$ | $8.945 \mathrm{E}-01$ | 2.693E-01 | 0.345 |
| 2004 | $8.086 \mathrm{E}-01$ | $1.596 \mathrm{E}-02$ | 1.97\% | $4.880 \mathrm{E}-01$ | $1.022 \mathrm{E}+00$ | $6.277 \mathrm{E}-01$ | $9.232 \mathrm{E}-01$ | 2.955E-01 | 0.366 |
| 2005 | 9.346E-01 | $1.626 \mathrm{E}-02$ | 1.74\% | $5.632 \mathrm{E}-01$ | $1.120 \mathrm{E}+00$ | $6.894 \mathrm{E}-01$ | $1.030 \mathrm{E}+00$ | 3.409E-01 | 0.365 |
| 2006 | $1.042 \mathrm{E}+00$ | $1.229 \mathrm{E}-02$ | 1.18\% | $6.258 \mathrm{E}-01$ | $1.218 \mathrm{E}+00$ | $7.677 \mathrm{E}-01$ | $1.123 \mathrm{E}+00$ | $3.550 \mathrm{E}-01$ | 0.341 |
| 2007 | $1.129 \mathrm{E}+00$ | $5.830 \mathrm{E}-03$ | 0.52\% | $6.787 \mathrm{E}-01$ | $1.294 \mathrm{E}+00$ | $8.459 \mathrm{E}-01$ | $1.200 \mathrm{E}+00$ | 3.542E-01 | 0.314 |
| 2008 | $1.195 \mathrm{E}+00$ | -7.869E-04 | -0.07\% | $7.944 \mathrm{E}-01$ | $1.350 \mathrm{E}+00$ | $9.237 \mathrm{E}-01$ | $1.265 \mathrm{E}+00$ | $3.410 \mathrm{E}-01$ | 0.285 |
| 2009 | $1.243 \mathrm{E}+00$ | -6.079E-03 | -0.49\% | $8.598 \mathrm{E}-01$ | $1.386 \mathrm{E}+00$ | $1.002 \mathrm{E}+00$ | $1.310 \mathrm{E}+00$ | $3.076 \mathrm{E}-01$ | 0.248 |
| 2010 | $1.277 \mathrm{E}+00$ | -9.577E-03 | -0.75\% | $8.999 \mathrm{E}-01$ | $1.405 \mathrm{E}+00$ | $1.043 \mathrm{E}+00$ | $1.336 \mathrm{E}+00$ | $2.937 \mathrm{E}-01$ | 0.230 |
| 2011 | $1.301 \mathrm{E}+00$ | -1.143E-02 | -0.88\% | $9.627 \mathrm{E}-01$ | $1.426 \mathrm{E}+00$ | $1.096 \mathrm{E}+00$ | $1.363 \mathrm{E}+00$ | $2.668 \mathrm{E}-01$ | 0.205 |
| 2012 | $1.317 \mathrm{E}+00$ | -1.205E-02 | -0.91\% | $1.004 \mathrm{E}+00$ | $1.437 \mathrm{E}+00$ | $1.130 \mathrm{E}+00$ | $1.380 \mathrm{E}+00$ | $2.501 \mathrm{E}-01$ | 0.190 |
| 2013 | $1.328 \mathrm{E}+00$ | -1.183E-02 | -0.89\% | $1.031 \mathrm{E}+00$ | $1.444 \mathrm{E}+00$ | $1.159 \mathrm{E}+00$ | $1.391 \mathrm{E}+00$ | 2.329E-01 | 0.175 |

NOTE: Confidence intervals are approximate.
At least 500 to 1000 trials are recommended when estimating confidence intervals. Results from ASPICP.EXE, version 3.10

Greenland halibut
halibut XIV
and
V
Page 2
Project $2 / 3 F_{\text {pa }}$
TRAJECTORY OF RELATIVE FISHING MORTALITY RATE F/F $\mathrm{F}_{\mathrm{MSY}}$ (BOOTSTRAPPED)


Table 6.7.3 (Cont'd)

| 1998 | $1.062 \mathrm{E}+00$ | $1.342 \mathrm{E}-02$ | 1.26\% | $7.887 \mathrm{E}-01$ | $1.313 \mathrm{E}+00$ | $9.145 \mathrm{E}-01$ | $1.200 \mathrm{E}+00$ | $2.856 \mathrm{E}-01$ | 0.269 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 9.372E-01 | 7.766E-03 | 0.83\% | $7.254 \mathrm{E}-01$ | $1.164 \mathrm{E}+00$ | $8.266 \mathrm{E}-01$ | $1.064 \mathrm{E}+00$ | $2.369 \mathrm{E}-01$ | 0.253 |
| 2000 | $1.111 \mathrm{E}+00$ | $4.234 \mathrm{E}-03$ | 0.38\% | 8.847E-01 | $1.398 \mathrm{E}+00$ | 9.872E-01 | $1.269 \mathrm{E}+00$ | $2.816 \mathrm{E}-01$ | 0.253 |
| 2001 | $1.109 \mathrm{E}+00$ | $2.970 \mathrm{E}-03$ | 0.27\% | $9.116 \mathrm{E}-01$ | $1.449 \mathrm{E}+00$ | $9.965 \mathrm{E}-01$ | $1.298 \mathrm{E}+00$ | $3.012 \mathrm{E}-01$ | 0.272 |
| 2002 | $1.083 \mathrm{E}+00$ | 1.980E-03 | $0.18 \%$ | $9.032 \mathrm{E}-01$ | $1.471 \mathrm{E}+00$ | $9.775 \mathrm{E}-01$ | $1.299 \mathrm{E}+00$ | $3.218 \mathrm{E}-01$ | 0.297 |
| 2003 | $1.086 \mathrm{E}+00$ | $3.430 \mathrm{E}-03$ | $0.32 \%$ | $9.047 \mathrm{E}-01$ | $1.563 \mathrm{E}+00$ | $9.891 \mathrm{E}-01$ | $1.381 \mathrm{E}+00$ | $3.924 \mathrm{E}-01$ | 0.361 |
| 2004 | $6.497 \mathrm{E}-01$ | $1.188 \mathrm{E}-03$ | 0.18\% | $5.419 \mathrm{E}-01$ | $8.828 \mathrm{E}-01$ | $5.865 \mathrm{E}-01$ | $7.796 \mathrm{E}-01$ | $1.931 \mathrm{E}-01$ | 0.297 |
| 2005 | $6.497 \mathrm{E}-01$ | $1.188 \mathrm{E}-03$ | 0.18\% | $5.419 \mathrm{E}-01$ | $8.828 \mathrm{E}-01$ | $5.865 \mathrm{E}-01$ | $7.796 \mathrm{E}-01$ | $1.931 \mathrm{E}-01$ | 0.297 |
| 2006 | $6.497 \mathrm{E}-01$ | $1.188 \mathrm{E}-03$ | 0.18\% | $5.419 \mathrm{E}-01$ | $8.828 \mathrm{E}-01$ | $5.865 \mathrm{E}-01$ | $7.796 \mathrm{E}-01$ | $1.931 \mathrm{E}-01$ | 0.297 |
| 2007 | $6.497 \mathrm{E}-01$ | $1.188 \mathrm{E}-03$ | 0.18\% | $5.419 \mathrm{E}-01$ | 8.828E-01 | $5.865 \mathrm{E}-01$ | $7.796 \mathrm{E}-01$ | $1.931 \mathrm{E}-01$ | 0.297 |
| 2008 | $6.497 \mathrm{E}-01$ | $1.188 \mathrm{E}-03$ | $0.18 \%$ | $5.419 \mathrm{E}-01$ | 8.828E-01 | $5.865 \mathrm{E}-01$ | $7.796 \mathrm{E}-01$ | $1.931 \mathrm{E}-01$ | 0.297 |
| 2009 | $6.497 \mathrm{E}-01$ | $1.188 \mathrm{E}-03$ | 0.18\% | $5.419 \mathrm{E}-01$ | 8.828E-01 | $5.865 \mathrm{E}-01$ | $7.796 \mathrm{E}-01$ | $1.931 \mathrm{E}-01$ | 0.297 |
| 2010 | $6.497 \mathrm{E}-01$ | $1.188 \mathrm{E}-03$ | 0.18\% | $5.419 \mathrm{E}-01$ | $8.828 \mathrm{E}-01$ | $5.865 \mathrm{E}-01$ | $7.796 \mathrm{E}-01$ | $1.931 \mathrm{E}-01$ | 0.297 |
| 2011 | $6.497 \mathrm{E}-01$ | $1.188 \mathrm{E}-03$ | 0.18\% | $5.419 \mathrm{E}-01$ | 8.828E-01 | $5.865 \mathrm{E}-01$ | $7.796 \mathrm{E}-01$ | $1.931 \mathrm{E}-01$ | 0.297 |
| 2012 | $6.497 \mathrm{E}-01$ | $1.188 \mathrm{E}-03$ | 0.18\% | $5.419 \mathrm{E}-01$ | 8.828E-01 | $5.865 \mathrm{E}-01$ | $7.796 \mathrm{E}-01$ | $1.931 \mathrm{E}-01$ | 0.297 |


| 2003 | $3.000 \mathrm{E}+04$ | $0.000 \mathrm{E}+00$ | 0.00\% | $3.000 \mathrm{E}+04$ | $3.000 \mathrm{E}+04$ | $3.000 \mathrm{E}+04$ | $3.000 \mathrm{E}+04$ | $0.000 \mathrm{E}+00$ | 0.000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | $1.972 \mathrm{E}+04$ | $1.756 \mathrm{E}+02$ | 0.89\% | $1.811 \mathrm{E}+04$ | $2.212 \mathrm{E}+04$ | $1.898 \mathrm{E}+04$ | $2.091 \mathrm{E}+04$ | 1.923E+03 | 0.098 |
| 2005 | $2.237 \mathrm{E}+04$ | $3.185 \mathrm{E}+02$ | 1.42\% | $1.967 \mathrm{E}+04$ | $2.626 \mathrm{E}+04$ | $2.099 \mathrm{E}+04$ | $2.427 \mathrm{E}+04$ | $3.282 \mathrm{E}+03$ | 0.147 |
| 2006 | $2.457 \mathrm{E}+04$ | $3.927 \mathrm{E}+02$ | 1.60\% | $2.106 \mathrm{E}+04$ | $2.928 \mathrm{E}+04$ | $2.270 \mathrm{E}+04$ | $2.715 \mathrm{E}+04$ | $4.451 \mathrm{E}+03$ | 0.181 |
| 2007 | $2.629 \mathrm{E}+04$ | $4.289 \mathrm{E}+02$ | 1.63\% | $2.227 \mathrm{E}+04$ | $3.228 \mathrm{E}+04$ | $2.408 \mathrm{E}+04$ | $2.937 \mathrm{E}+04$ | $5.290 \mathrm{E}+03$ | 0.201 |
| 2008 | $2.757 \mathrm{E}+04$ | $4.631 \mathrm{E}+02$ | 1.68\% | $2.320 \mathrm{E}+04$ | $3.472 \mathrm{E}+04$ | $2.507 \mathrm{E}+04$ | $3.092 \mathrm{E}+04$ | $5.858 \mathrm{E}+03$ | 0.212 |
| 2009 | $2.849 \mathrm{E}+04$ | $5.151 \mathrm{E}+02$ | 1.81\% | $2.375 \mathrm{E}+04$ | $3.630 \mathrm{E}+04$ | $2.578 \mathrm{E}+04$ | $3.212 \mathrm{E}+04$ | $6.340 \mathrm{E}+03$ | 0.223 |
| 2010 | $2.914 \mathrm{E}+04$ | $5.892 \mathrm{E}+02$ | 2.02\% | $2.405 \mathrm{E}+04$ | $3.740 \mathrm{E}+04$ | $2.630 \mathrm{E}+04$ | $3.290 \mathrm{E}+04$ | $6.604 \mathrm{E}+03$ | 0.227 |
| 2011 | $2.958 \mathrm{E}+04$ | $6.814 \mathrm{E}+02$ | 2.30\% | $2.449 \mathrm{E}+04$ | $3.823 \mathrm{E}+04$ | $2.682 \mathrm{E}+04$ | $3.368 \mathrm{E}+04$ | $6.857 \mathrm{E}+03$ | 0.232 |
| 2012 | $2.989 \mathrm{E}+04$ | 7.849E+02 | 2.63\% | $2.472 \mathrm{E}+04$ | $3.889 \mathrm{E}+04$ | $2.708 \mathrm{E}+04$ | $3.429 \mathrm{E}+04$ | $7.210 \mathrm{E}+03$ | 0.241 |

NOTE: Confidence intervals are approximate.
At least 500 to 1000 trials are recommended when estimating confidence intervals.
TRAJECTORY OF ABSOLUTE BIOMASS (BOOTSTRAPPED)

| Year | Point estimate | Estimated bias | Relative bias | Approx 80\% lower CL | $\begin{aligned} & \text { Approx } 80 \% \\ & \text { upper CL } \end{aligned}$ | Approx 50\% lower CL | Approx 50\% upper CL | Interquartile range | Relative <br> IQ range |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | $1.724 \mathrm{E}+05$ | $2.555 \mathrm{E}+04$ | 14.82\% | $9.578 \mathrm{E}+04$ | $3.216 \mathrm{E}+05$ | $1.289 \mathrm{E}+05$ | $2.400 \mathrm{E}+05$ | $1.111 \mathrm{E}+05$ | 0.644 |
| 1986 | $1.671 \mathrm{E}+05$ | $1.164 \mathrm{E}+04$ | 6.97\% | $1.078 \mathrm{E}+05$ | $2.854 \mathrm{E}+05$ | $1.347 \mathrm{E}+05$ | $2.182 \mathrm{E}+05$ | $8.352 \mathrm{E}+04$ | 0.500 |
| 1987 | $1.623 \mathrm{E}+05$ | $6.482 \mathrm{E}+03$ | 3.99\% | $1.149 \mathrm{E}+05$ | $2.641 \mathrm{E}+05$ | $1.361 \mathrm{E}+05$ | $2.011 \mathrm{E}+05$ | $6.494 \mathrm{E}+04$ | 0.400 |
| 1988 | $1.464 \mathrm{E}+05$ | $4.621 \mathrm{E}+03$ | 3.16\% | $1.067 \mathrm{E}+05$ | $2.390 \mathrm{E}+05$ | $1.247 \mathrm{E}+05$ | $1.813 \mathrm{E}+05$ | $5.650 \mathrm{E}+04$ | 0.386 |
| 1989 | $1.287 \mathrm{E}+05$ | $4.081 \mathrm{E}+03$ | 3.17\% | $9.680 \mathrm{E}+04$ | $2.139 \mathrm{E}+05$ | $1.108 \mathrm{E}+05$ | $1.626 \mathrm{E}+05$ | $5.182 \mathrm{E}+04$ | 0.403 |
| 1990 | $1.019 \mathrm{E}+05$ | $3.915 \mathrm{E}+03$ | 3.84\% | $7.250 \mathrm{E}+04$ | $1.773 \mathrm{E}+05$ | $8.553 \mathrm{E}+04$ | $1.317 \mathrm{E}+05$ | $4.613 \mathrm{E}+04$ | 0.453 |
| 1991 | $9.674 \mathrm{E}+04$ | $3.811 \mathrm{E}+03$ | 3.94\% | $6.974 \mathrm{E}+04$ | $1.662 \mathrm{E}+05$ | $8.162 \mathrm{E}+04$ | $1.248 \mathrm{E}+05$ | $4.315 \mathrm{E}+04$ | 0.446 |
| 1992 | $9.249 \mathrm{E}+04$ | $3.848 \mathrm{E}+03$ | 4.16\% | $6.705 \mathrm{E}+04$ | $1.566 \mathrm{E}+05$ | $7.803 \mathrm{E}+04$ | $1.186 \mathrm{E}+05$ | $4.055 \mathrm{E}+04$ | 0.438 |
| 1993 | $9.035 \mathrm{E}+04$ | $4.052 \mathrm{E}+03$ | 4.49\% | $6.700 \mathrm{E}+04$ | $1.507 \mathrm{E}+05$ | $7.737 \mathrm{E}+04$ | $1.160 \mathrm{E}+05$ | $3.864 \mathrm{E}+04$ | 0.428 |
| 1994 | $8.165 \mathrm{E}+04$ | $4.369 \mathrm{E}+03$ | 5.35\% | $5.980 \mathrm{E}+04$ | $1.377 \mathrm{E}+05$ | $6.891 \mathrm{E}+04$ | $1.061 \mathrm{E}+05$ | $3.718 \mathrm{E}+04$ | 0.455 |
| 1995 | $7.578 \mathrm{E}+04$ | $4.738 \mathrm{E}+03$ | 6.25\% | $5.494 \mathrm{E}+04$ | $1.296 \mathrm{E}+05$ | $6.337 \mathrm{E}+04$ | $9.651 \mathrm{E}+04$ | $3.314 \mathrm{E}+04$ | 0.437 |
| 1996 | $6.957 \mathrm{E}+04$ | $5.139 \mathrm{E}+03$ | 7.39\% | $4.922 \mathrm{E}+04$ | $1.219 \mathrm{E}+05$ | $5.731 \mathrm{E}+04$ | $8.956 \mathrm{E}+04$ | $3.225 \mathrm{E}+04$ | 0.464 |
| 1997 | $6.213 \mathrm{E}+04$ | $5.448 \mathrm{E}+03$ | 8.77\% | $4.156 \mathrm{E}+04$ | $1.157 \mathrm{E}+05$ | $5.022 \mathrm{E}+04$ | $8.474 \mathrm{E}+04$ | $3.452 \mathrm{E}+04$ | 0.556 |
| 1998 | $5.946 \mathrm{E}+04$ | $5.510 \mathrm{E}+03$ | 9.27\% | $3.617 \mathrm{E}+04$ | $1.128 \mathrm{E}+05$ | $4.603 \mathrm{E}+04$ | $8.067 \mathrm{E}+04$ | $3.464 \mathrm{E}+04$ | 0.583 |
| 1999 | $6.686 \mathrm{E}+04$ | $5.356 \mathrm{E}+03$ | 8.01\% | $4.140 \mathrm{E}+04$ | $1.215 \mathrm{E}+05$ | $5.228 \mathrm{E}+04$ | $8.952 \mathrm{E}+04$ | $3.724 \mathrm{E}+04$ | 0.557 |
| 2000 | $7.635 \mathrm{E}+04$ | $5.155 \mathrm{E}+03$ | 6.75\% | $4.871 \mathrm{E}+04$ | $1.263 \mathrm{E}+05$ | $6.056 \mathrm{E}+04$ | $9.874 \mathrm{E}+04$ | $3.818 \mathrm{E}+04$ | 0.500 |
| 2001 | $8.116 \mathrm{E}+04$ | $4.970 \mathrm{E}+03$ | 6.12\% | $5.127 \mathrm{E}+04$ | $1.254 \mathrm{E}+05$ | $6.381 \mathrm{E}+04$ | $1.024 \mathrm{E}+05$ | $3.860 \mathrm{E}+04$ | 0.476 |
| 2002 | $8.531 \mathrm{E}+04$ | $4.795 \mathrm{E}+03$ | 5.62\% | $5.468 \mathrm{E}+04$ | $1.269 \mathrm{E}+05$ | $6.767 \mathrm{E}+04$ | $1.046 \mathrm{E}+05$ | $3.696 \mathrm{E}+04$ | 0.433 |
| 2003 | $8.941 \mathrm{E}+04$ | $4.642 \mathrm{E}+03$ | 5.19\% | $5.830 \mathrm{E}+04$ | $1.283 \mathrm{E}+05$ | $7.125 \mathrm{E}+04$ | $1.071 \mathrm{E}+05$ | $3.587 \mathrm{E}+04$ | 0.401 |
| 2004 | 9.272E+04 | $4.498 \mathrm{E}+03$ | 4.85\% | $6.042 \mathrm{E}+04$ | $1.284 \mathrm{E}+05$ | $7.392 \mathrm{E}+04$ | $1.093 \mathrm{E}+05$ | $3.542 \mathrm{E}+04$ | 0.382 |
| 2005 | $1.072 \mathrm{E}+05$ | $4.140 \mathrm{E}+03$ | 3.86\% | $7.454 \mathrm{E}+04$ | $1.423 \mathrm{E}+05$ | $9.028 \mathrm{E}+04$ | $1.236 \mathrm{E}+05$ | $3.337 \mathrm{E}+04$ | 0.311 |
| 2006 | $1.195 \mathrm{E}+05$ | $3.598 \mathrm{E}+03$ | 3.01\% | $8.821 \mathrm{E}+04$ | $1.570 \mathrm{E}+05$ | $1.040 \mathrm{E}+05$ | $1.378 \mathrm{E}+05$ | $3.381 \mathrm{E}+04$ | 0.283 |
| 2007 | $1.295 \mathrm{E}+05$ | $3.056 \mathrm{E}+03$ | 2.36\% | $9.682 \mathrm{E}+04$ | $1.670 \mathrm{E}+05$ | $1.141 \mathrm{E}+05$ | $1.471 \mathrm{E}+05$ | $3.299 \mathrm{E}+04$ | 0.255 |
| 2008 | $1.370 \mathrm{E}+05$ | $2.691 \mathrm{E}+03$ | 1.96\% | $1.054 \mathrm{E}+05$ | $1.771 \mathrm{E}+05$ | $1.218 \mathrm{E}+05$ | $1.557 \mathrm{E}+05$ | $3.385 \mathrm{E}+04$ | 0.247 |
| 2009 | $1.425 \mathrm{E}+05$ | $2.581 \mathrm{E}+03$ | 1.81\% | $1.097 \mathrm{E}+05$ | $1.862 \mathrm{E}+05$ | $1.255 \mathrm{E}+05$ | $1.615 \mathrm{E}+05$ | $3.603 \mathrm{E}+04$ | 0.253 |
| 2010 | $1.464 \mathrm{E}+05$ | $2.716 \mathrm{E}+03$ | 1.85\% | $1.127 \mathrm{E}+05$ | 1.950E+05 | $1.287 \mathrm{E}+05$ | $1.668 \mathrm{E}+05$ | $3.814 \mathrm{E}+04$ | 0.260 |
| 2011 | $1.492 \mathrm{E}+05$ | $3.037 \mathrm{E}+03$ | 2.04\% | $1.163 \mathrm{E}+05$ | $1.998 \mathrm{E}+05$ | $1.308 \mathrm{E}+05$ | $1.707 \mathrm{E}+05$ | $3.995 \mathrm{E}+04$ | 0.268 |
| 2012 | $1.510 \mathrm{E}+05$ | $3.481 \mathrm{E}+03$ | 2.30\% | $1.184 \mathrm{E}+05$ | $2.052 \mathrm{E}+05$ | $1.327 \mathrm{E}+05$ | $1.735 \mathrm{E}+05$ | $4.074 \mathrm{E}+04$ | 0.270 |
| 2013 | $1.523 \mathrm{E}+05$ | $3.992 \mathrm{E}+03$ | 2.62\% | $1.193 \mathrm{E}+05$ | $2.083 \mathrm{E}+05$ | $1.338 \mathrm{E}+05$ | $1.754 \mathrm{E}+05$ | $4.161 \mathrm{E}+04$ | 0.273 |

NOTE: Confidence intervals are approximate.
At least 500 to 1000 trials are recommended when estimating confidence intervals.
TRAJECTORY OF ABSOLUTE FISHING MORTALITY RATE (BOOTSTRAPPED)

| Year | Point estimate | Estimated bias | Relative bias | $\begin{aligned} & \text { Approx 80\% } \\ & \text { lower CL } \end{aligned}$ | $\begin{aligned} & \text { Approx 80\% } \\ & \text { upper CL } \end{aligned}$ | $\begin{aligned} & \text { Approx 50\% } \\ & \text { lower CL } \end{aligned}$ | $\begin{aligned} & \text { Approx 50\% } \\ & \text { upper CL } \end{aligned}$ | Interquartile range | Relative <br> IQ range |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | $1.892 \mathrm{E}-01$ | $1.840 \mathrm{E}-02$ | 9.73\% | $1.044 \mathrm{E}-01$ | $3.100 \mathrm{E}-01$ | $1.386 \mathrm{E}-01$ | $2.418 \mathrm{E}-01$ | $1.032 \mathrm{E}-01$ | 0.545 |
| 1986 | $2.005 \mathrm{E}-01$ | $1.559 \mathrm{E}-02$ | 7.78\% | $1.200 \mathrm{E}-01$ | $2.969 \mathrm{E}-01$ | $1.580 \mathrm{E}-01$ | $2.440 \mathrm{E}-01$ | 8.599E-02 | 0.429 |
| 1987 | $3.033 \mathrm{E}-01$ | $2.004 \mathrm{E}-02$ | 6.61\% | $1.835 \mathrm{E}-01$ | $4.158 \mathrm{E}-01$ | $2.436 \mathrm{E}-01$ | $3.580 \mathrm{E}-01$ | $1.144 \mathrm{E}-01$ | 0.377 |
| 1988 | $3.736 \mathrm{E}-01$ | 2.322E-02 | 6.22\% | $2.253 \mathrm{E}-01$ | $5.089 \mathrm{E}-01$ | $3.000 \mathrm{E}-01$ | $4.350 \mathrm{E}-01$ | $1.349 \mathrm{E}-01$ | 0.361 |
| 1989 | $5.380 \mathrm{E}-01$ | 3.891E-02 | 7.23\% | $3.149 \mathrm{E}-01$ | $7.349 \mathrm{E}-01$ | $4.198 \mathrm{E}-01$ | $6.301 \mathrm{E}-01$ | 2.103E-01 | 0.391 |
| 1990 | $3.964 \mathrm{E}-01$ | $3.212 \mathrm{E}-02$ | 8.10\% | $2.288 \mathrm{E}-01$ | $5.531 \mathrm{E}-01$ | $3.064 \mathrm{E}-01$ | $4.701 \mathrm{E}-01$ | $1.637 \mathrm{E}-01$ | 0.413 |
| 1991 | $4.015 \mathrm{E}-01$ | $3.048 \mathrm{E}-02$ | 7.59\% | $2.354 \mathrm{E}-01$ | $5.562 \mathrm{E}-01$ | $3.123 \mathrm{E}-01$ | $4.756 \mathrm{E}-01$ | $1.633 \mathrm{E}-01$ | 0.407 |
| 1992 | $3.884 \mathrm{E}-01$ | $2.584 \mathrm{E}-02$ | 6.65\% | $2.309 \mathrm{E}-01$ | $5.293 \mathrm{E}-01$ | $3.028 \mathrm{E}-01$ | $4.577 \mathrm{E}-01$ | $1.549 \mathrm{E}-01$ | 0.399 |
| 1993 | $4.810 \mathrm{E}-01$ | $2.940 \mathrm{E}-02$ | 6.11\% | $2.852 \mathrm{E}-01$ | $6.527 \mathrm{E}-01$ | $3.716 \mathrm{E}-01$ | $5.618 \mathrm{E}-01$ | $1.903 \mathrm{E}-01$ | 0.396 |
| 1994 | $4.733 \mathrm{E}-01$ | $2.851 \mathrm{E}-02$ | 6.02\% | $2.781 \mathrm{E}-01$ | $6.495 \mathrm{E}-01$ | $3.718 \mathrm{E}-01$ | $5.646 \mathrm{E}-01$ | $1.928 \mathrm{E}-01$ | 0.407 |
| 1995 | $5.003 \mathrm{E}-01$ | $3.045 \mathrm{E}-02$ | 6.09\% | $2.886 \mathrm{E}-01$ | $6.997 \mathrm{E}-01$ | $3.915 \mathrm{E}-01$ | $6.017 \mathrm{E}-01$ | 2.102E-01 | 0.420 |
| 1996 | $5.460 \mathrm{E}-01$ | $4.036 \mathrm{E}-02$ | 7.39\% | $3.020 \mathrm{E}-01$ | $7.924 \mathrm{E}-01$ | $4.120 \mathrm{E}-01$ | $6.700 \mathrm{E}-01$ | $2.580 \mathrm{E}-01$ | 0.472 |
| 1997 | $4.897 \mathrm{E}-01$ | 5.129E-02 | 10.47\% | $2.576 \mathrm{E}-01$ | $7.667 \mathrm{E}-01$ | $3.543 \mathrm{E}-01$ | $6.156 \mathrm{E}-01$ | $2.613 \mathrm{E}-01$ | 0.534 |
| 1998 | 3.222E-01 | $3.930 \mathrm{E}-02$ | 12.20\% | $1.735 \mathrm{E}-01$ | $5.285 \mathrm{E}-01$ | $2.383 \mathrm{E}-01$ | $4.144 \mathrm{E}-01$ | $1.762 \mathrm{E}-01$ | 0.547 |
| 1999 | $2.842 \mathrm{E}-01$ | $3.066 \mathrm{E}-02$ | 10.79\% | $1.669 \mathrm{E}-01$ | $4.570 \mathrm{E}-01$ | $2.180 \mathrm{E}-01$ | $3.638 \mathrm{E}-01$ | $1.458 \mathrm{E}-01$ | 0.513 |
| 2000 | $3.371 \mathrm{E}-01$ | $3.160 \mathrm{E}-02$ | 9.37\% | $2.068 \mathrm{E}-01$ | $5.263 \mathrm{E}-01$ | $2.648 \mathrm{E}-01$ | $4.294 \mathrm{E}-01$ | $1.646 \mathrm{E}-01$ | 0.488 |
| 2001 | $3.364 \mathrm{E}-01$ | $2.887 \mathrm{E}-02$ | 8.58\% | $2.124 \mathrm{E}-01$ | $5.192 \mathrm{E}-01$ | $2.666 \mathrm{E}-01$ | $4.198 \mathrm{E}-01$ | $1.532 \mathrm{E}-01$ | 0.455 |
| 2002 | $3.284 \mathrm{E}-01$ | $2.541 \mathrm{E}-02$ | 7.74\% | $2.263 \mathrm{E}-01$ | $5.060 \mathrm{E}-01$ | $2.719 \mathrm{E}-01$ | $4.162 \mathrm{E}-01$ | $1.442 \mathrm{E}-01$ | 0.439 |
| 2003 | $3.292 \mathrm{E}-01$ | $2.340 \mathrm{E}-02$ | 7.11\% | $2.344 \mathrm{E}-01$ | $5.017 \mathrm{E}-01$ | $2.783 \mathrm{E}-01$ | $4.097 \mathrm{E}-01$ | $1.314 \mathrm{E}-01$ | 0.399 |
| 2004 | $1.970 \mathrm{E}-01$ | $1.525 \mathrm{E}-02$ | 7.74\% | $1.358 \mathrm{E}-01$ | $3.036 \mathrm{E}-01$ | $1.632 \mathrm{E}-01$ | $2.497 \mathrm{E}-01$ | $8.654 \mathrm{E}-02$ | 0.439 |
| 2005 | $1.970 \mathrm{E}-01$ | $1.525 \mathrm{E}-02$ | 7.74\% | $1.358 \mathrm{E}-01$ | $3.036 \mathrm{E}-01$ | $1.632 \mathrm{E}-01$ | $2.497 \mathrm{E}-01$ | $8.654 \mathrm{E}-02$ | 0.439 |
| 2006 | $1.970 \mathrm{E}-01$ | $1.525 \mathrm{E}-02$ | 7.74\% | $1.358 \mathrm{E}-01$ | $3.036 \mathrm{E}-01$ | $1.632 \mathrm{E}-01$ | $2.497 \mathrm{E}-01$ | $8.654 \mathrm{E}-02$ | 0.439 |
| 2007 | $1.970 \mathrm{E}-01$ | $1.525 \mathrm{E}-02$ | 7.74\% | $1.358 \mathrm{E}-01$ | $3.036 \mathrm{E}-01$ | $1.632 \mathrm{E}-01$ | $2.497 \mathrm{E}-01$ | 8.654E-02 | 0.439 |
| 2008 | $1.970 \mathrm{E}-01$ | $1.525 \mathrm{E}-02$ | 7.74\% | $1.358 \mathrm{E}-01$ | $3.036 \mathrm{E}-01$ | $1.632 \mathrm{E}-01$ | $2.497 \mathrm{E}-01$ | $8.654 \mathrm{E}-02$ | 0.439 |
| 2009 | $1.970 \mathrm{E}-01$ | $1.525 \mathrm{E}-02$ | 7.74\% | $1.358 \mathrm{E}-01$ | $3.036 \mathrm{E}-01$ | $1.632 \mathrm{E}-01$ | $2.497 \mathrm{E}-01$ | $8.654 \mathrm{E}-02$ | 0.439 |
| 2010 | $1.970 \mathrm{E}-01$ | $1.525 \mathrm{E}-02$ | 7.74\% | $1.358 \mathrm{E}-01$ | $3.036 \mathrm{E}-01$ | $1.632 \mathrm{E}-01$ | $2.497 \mathrm{E}-01$ | $8.654 \mathrm{E}-02$ | 0.439 |
| 2011 | $1.970 \mathrm{E}-01$ | $1.525 \mathrm{E}-02$ | 7.74\% | $1.358 \mathrm{E}-01$ | $3.036 \mathrm{E}-01$ | $1.632 \mathrm{E}-01$ | $2.497 \mathrm{E}-01$ | $8.654 \mathrm{E}-02$ | 0.439 |
| 2012 | $1.970 \mathrm{E}-01$ | $1.525 \mathrm{E}-02$ | 7.74\% | $1.358 \mathrm{E}-01$ | $3.036 \mathrm{E}-01$ | $1.632 \mathrm{E}-01$ | $2.497 \mathrm{E}-01$ | 8.654E-02 | 0.439 |



Figure 6.1.1 Landings of Greenland halibut in Divisions Va, Vb and Subarea XIV. As the landings within Icelandic waters, since 1976, have not officially been separated and reported according to the defined ICES statistical areas, they are set under area Va by the North Western Working Group.
Div. XIVb

Div. Va

Div. Vb


Figure 6.2.1 Standardised CPUE series from fleets in Divisions XIVb, Va and Vb with indication of 95\% confidence limits.



Figure 6.6.1 Greenland halibut in Icelandic fall groundfish survey a) biomass indices of lengths larger than indicated and b) abundance indices by lengths smaller than indicated.

## Greenland Survey East Greenland



Figure 6.6.2 Estimated trawlable biomass in Division XIVb from the Greenland deep-water trawl survey with $95 \%$ confidence limits indicated.


Figure 6.6.3 Age distribution in Greenland deep-water survey in Div. XIVb in 1999, 2000 and 2002.


Figure 6.7.2 $1 \quad$ Observed and predicted CPUE's. Upper: Icelandic groundfish survey, Lower: Icelandic trawler CPUE


Figure 6.7.2.2 Output from ASPIC (Table 6.7.2.1.) with B/B $\mathbf{B}_{\text {MSY }}$ and $\mathrm{F} / \mathbf{F}_{\mathrm{MSY}}$.

30 kt in 2003 and 2/3Fmsy 2004 onwards


30 kt in 2003 and Fsq 2004 onwards


30 kt in 2003 and onwards


Figure 6.7.3 Biomass $\left(\mathrm{B} / \mathbf{B}_{\mathrm{MSY}}\right)$ trajectories under different options as derived from ASPIC-P.

Species of the genus Sebastes are common and widely distributed in the North Atlantic. They are found off the coast of Great Britain, along Norway and Spitzbergen, in the Barents Sea, off the Faroe Islands, Iceland, East and West Greenland, and along the east coast of North America from Baffin Island to Cape Cod. All Sebastes species are viviparous. The extrusion of the larvae takes place in late winter-late spring/early summer, but copulation occurs in autumn-early winter.

There are three species of redfish commercially exploited in ICES Subareas V, VI, XII, and XIV, S. marinus, $S$. mentella, and $S$. viviparus. The last one has only been of a minor commercial value in Icelandic waters and is exploited in 2 small areas south of Iceland at depths of 150-250 m. The landings of S. viviparus decreased from 1,160 t in 1994 to 20 t in 2002.

### 7.1 Problems regarding stock identity of S. mentella

The existence of more than one stock of $S$. mentella in the area has been discussed in recent years. Historically, $S$. mentella was fished on the continental shelves and slopes of the Faroe Islands, Iceland, and East Greenland and been considered as one stock. A new pelagic fishery started in the open Irminger Sea in 1982, primarily fishing in waters shallower than 500 m . In 1992, the Study Group on Redfish Stocks distinguished between these types as deep-sea $S$. mentella (shelf redfish) and oceanic $S$. mentella (Irminger Sea redfish). In the early 1990's, the pelagic fishery in the open Irminger Sea moved to layers deeper than 500 m . Some researchers considered that the fish caught pelagically deeper than 500 m differed from the fish caught shallower than 500 m and resembled more to the deep-sea $S$. mentella living on the continental shelves and slopes. $S$. mentella living deeper than 500 m has been called "pelagic deep-sea $S$. mentella". Recently, the distribution of the pelagic $S$. mentella in the upper 500 m has extended significantly more southwest and into the NAFO Convention Areas compared to the early 1990's.

It is not known whether these types represent one stock or several biologically different stocks and different hypotheses have been put forward based on comprehensive studies on growth, maturity, morphometrics, parasites as natural tags, and genetic and fatty acid differentiation of the species:

- Single-stock hypothesis: All $S$. mentella from the Faroe Islands to the Grand Banks is one stock and is segregated according to age/size.
- Two-stock hypothesis: The $S$. mentella living on the shelves (deep-sea $S$. mentella) and those living in deeper pelagic waters of the Irminger Sea (pelagic deep-sea $S$. mentella) is one stock unit, which is separated from the oceanic $S$. mentella living in the upper layers of the Irminger Sea.
- Three-stock hypothesis: The three described components are biologically different stocks.

Despite a lot of effort by the WG, there is not a consensus within the WG regarding which hypothesis is the most likely one. Although the uncertainty regarding stock structure of $S$. mentella is great, extensive research have been done. Currently, several studies are ongoing to answer important questions regarding the biology, population structure, and abundance and demography of this highly migratory and straddling species.

### 7.2 Nominal catches and splitting of the landings into stocks

The official statistics reported to ICES do not divide catch by species/stocks (Tables 7.2.1-7.2.5). Information from various sources, for example, from samples taken from the catch in different fishing areas and information on products, are used to split catches into species and stocks. The technique and the data used for such splitting were described in the 1998 NWWG report.

### 7.3 Abundance and distribution of 0-group and juvenile redfish

Available data on the distribution of juvenile S. marinus indicate that the nursery grounds are located in Icelandic and Greenland waters. No nursery grounds have been found in Faroese waters. Studies indicate that considerable amounts of juvenile S. marinus of East Greenland is mixed with juvenile S. mentella (Magnússon et al. 1988; 1990, ICES CM 1998/G:3). The 1983 Redfish Study Group report (ICES CM 1983/G:3) and in Magnússon and Jóhannesson (1997) describes the distribution of 0 -group $S$. marinus off East Greenland. The nursery areas for $S$. marinus in Icelandic waters are found all around Iceland, but are mainly located west and north of the island at depths between 50 and 350 m
(ICES C.M.1983/G:3; Einarsson, 1960; Magnússon and Magnússon 1975; Pálsson et al. 1997). The migration of juveniles is along the north coast towards the most important fishing areas off the west coast.

Indices for 0 -group redfish in the Irminger Sea and at East Greenland areas were available from the Icelandic 0 -group surveys from 1970-1995. Thereafter, the survey was discontinued. Above or average year class strengths were observed in 1972, 1973-74, 1985-91, and in 1995.

Abundance and biomass indices of juvenile ( $<17 \mathrm{~cm}$ ) redfish (juveniles were only classified to the genus Sebastes spp. due to difficult identification) from the German annual groundfish survey, conducted on the continental shelf and slope of West and East Greenland down to 400 m, shows that juveniles were abundant in 1993 and 1995-1998 (Figure 7.3.1). The 1999-2002 survey results indicate low abundance and are similar to those observed in the late 1980s.

### 7.4 Discards and by-catch of small redfish in East and West Greenland

An offshore shrimp fishery with small-meshed trawl (44 mm in the codend) began in the early 1970s off the west coast of Greenland. This fishery expanded to the east coast in the beginning of the 1980s and was mainly conducted on the shallower part of the Dohrn Bank and on the continental shelf from $65^{\circ} \mathrm{N}$ to $60^{\circ} \mathrm{N}$. Observer samples from the Greenland Fishery Licence Control showed that redfish is by-catch in the shrimp fishery off Greenland. No information was available in recent years to quantify the by-catch and about the length distribution of the fish caught. Since the 1st October 2000, sorting grids have been mandatory to reduce by-catch, but the effect has not been documented.

### 7.5 Special Requests

There are several questions regarding stock structure, distribution, and fishery information of $S$. mentella in the area in the ToR for the Working Group. The following paragraphs deal with ToR $c, e$, and $f$ and special requests $b$ and $c$ from NEAFC. The WG also deals with these questions in some cases in more detail under different redfish Sections.

ToR c). Detailed descriptions of the fishery of different nations are given in Sections 8 for $S$. marinus, 9 for deep-sea $S$. mentella, and 10 for oceanic $S$. mentella, based on various working documents.

The fishery for oceanic $S$. mentella in ICES Subareas Va, XII, and XIV and in NAFO areas shows a persistent seasonal pattern in terms of geographical and depth distribution for the past five years (Figures 7.5.1-7.5.4). The main fishing occurs in the second and third quarter of the year. In the second quarter, the fishery takes place in the area east of $32^{\circ} \mathrm{W}$ and north of $61^{\circ} \mathrm{N}$ at depths deeper than 500 m . In the third quarter, the fleet moves towards the southwest to ICES Subarea XII and NAFO Convention areas and the depth of the hauls are in waters shallower than 500 m . There has traditionally been very little fishing activity from November until late March, and in 2002 no activity was reported during that time. The size of the fish caught in the southwest areas in the third quarter of the year is smaller than the fish caught in the northwest area in the second quarter (Figure 7.5.5). The fish caught in all seasons are sexually mature.

Based on the geographical and seasonal distribution of the oceanic $S$. mentella, catches in the Irminger Sea and adjacent waters in 2002 (Figures 7.5.1-7.5.4) it was concluded that the fishing pattern in 2002 was similar as it was in the past five years.

As has been reported in earlier reports of the Working Group, Iceland has classified its pelagic catches between oceanic and pelagic deep-sea redfish. Based on the samples, the results indicated that at depths shallower than $500-600 \mathrm{~m}$, the proportion "oceanic" is between $85-100 \%$, and the proportion deeper than 600 m between $0-20 \%$.

The WG acknowledge information on trawling depth as provided by some nations, but recommends that all nations provide depth information in accordance with the NEAFC logbook format.

ToR e) and NEAFC special requests b) and c). New data presented in various working documents presents results of different methods that were used to investigate the issue of stock structure. Result in one paper (WD9) suggests some difference of the "pelagic deep sea $S$. mentella" and the $S$. mentella caught in the demersal fishery on the slope, concluding that there was "no big exchange between redfish stocks distributed on the south-western slope of Iceland and in the pelagic sea". WD30 describes recent changes in the pelagic fishery, where fishing areas of the pelagic $S$. mentella and deep-sea $S$. mentella on the slope in Division Va are now closer to each other. For management purposes the Icelandic authorities have separated these fisheries with the so-called redfish line (see Chapter 9), but this may not reflect two biologically different stocks. WD 8 suggests that for conservation and rational exploitation of the pelagic redfish stock a single TAC should be kept. Based on limited expertise of the WG it was concluded that the information
presented did not justify a change in the perception of the stock structure in relation to the current way management advise is given.

There is consensus that NWWG is primarily an assessment group and does not have sufficient expertise to thoroughly review the scientific research of redfish stock identification. The methodological approaches include various genetic differentiation, morphometrics, parasitology, growth patterns, and trace element analyses. In light of this and in light of the EU project and other relevant data on redfish that is expected to commence this fall, the WG agreed to recommend that a separate ICES group with the appropriate expertise would review both existing and pending scientific material. This could either dealt with in a special study group or possibly within the current ICES Stock Identification Methods Working Group. The group should report to the NWWG meeting in 2004.

ToR f). Limited information is available for describing the distribution of the stock(s) in the area throughout the year and the information from the international trawl-acoustic survey in 2001 did not add much to the current knowledge. Information from various acoustic estimates in recent years only describes the distribution at one time of the year (June/July). Information from the fishery of various nations cannot be used alone as a description of the distribution. These sources are thus not considered adequate to describe the seasonal distribution of the various components. A new international acoustic-trawl survey will be conducted in June 2003 and a report will be available before the ACFM fall meeting in 2003. It is not likely that this survey will add to the current knowledge on the issue of ToR f.

Table 7.2.1 REDFISH. Nominal catches (tonnes) by countries, in Division Va 1996-2002, as officially reported to ICES.

| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | $2002^{*}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Faroe Islands | 309 | 242 | 280 | 255 |  |  |  |
| Germany | 233 | - | 284 | 428 | 513 | 844 | 467 |
| Iceland | 67,757 | 73,976 | 108,380 | 81,430 | 95,118 | 48,970 | 66,449 |
| Norway | 134 | - | - | 18 | $36^{*}$ | $26^{*}$ | 16 |
| UK (E/W/NI) | - | - | - | 542 | 734 | 1,037 | $\ldots$ |
| UK (Scotland) | - | - | - | 149 | 70 | 114 | $\ldots$ |
| United Kingdom |  |  |  |  |  |  | 704 |
| Total | 68,433 | 74,218 | 108,944 | 82,822 |  |  |  |
| ${ }^{*}$ Preliminary. |  |  |  |  |  |  |  |

Table 7.2.2 REDFISH. Nominal catches (tonnes) by countries, in Division Vb 1996-2002, as officially reported to ICES.

| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | $2002^{*}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Faroe Islands | 7,286 | 7,199 | 6,484 | 6,191 |  |  |  |
| France | 62 | 98 | $110^{*}$ |  | 250 | $178^{*}$ | 207 |
| Germany | 189 | 36 | - | 207 | 79 | 88 | 2 |
| Iceland | - | - | - | - | - | 54 | - |
| Ireland | - | - | - | - | - | 1 |  |
| Norway | 33 | - | 25 | 39 | 37 | $42^{*}$ | $24^{*}$ |
| Russia | - | - | - | 12 | - | 30 |  |
| UK (E/W/NI) | 40 | - | 4 | 15 | 111 | 92 | - |
| UK (Scotland) | 43 | 36 | 27 | 46 | 142 | 116 | $\ldots$ |
| United Kingdom |  |  |  |  |  |  | 409 |
| Total | 7,653 |  |  |  |  |  |  |
| Preliminary. |  |  |  |  |  |  |  |

Table 7.2.3 REDFISH. Nominal catches (tonnes) by countries, in Division VI 1996-2002, as officially reported to ICES.

| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | $2002^{*}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Estonia | - | - | - | - | - | + |  |
| Faroe Islands | - | 12 | - | 44 |  |  |  |
| France | 489 | 395 | $297^{*}$ |  | 269 | $210^{*}$ | 96 |
| Germany | 9 | 1 | 1 | + | + | 1 | - |
| Ireland | - | 10 | 10 | 34 | 54 | 47 |  |
| Norway | 7 | 6 | 3 | 8 | $11^{*}$ | $5^{*}$ | - |
| Portugal | - | - | 1 | - | - | - | - |
| Russia | - | - | - | 243 | 461 | 88 | 19 |
| Spain | - | - | - | 38 | 16 | 4 |  |
| UK (E/W/NI) | 54 | 19 | 12 | 4 | 20 | 44 | $\ldots$ |
| UK (Scotland) | 603 | 518 | 364 | 762 | 405 | 485 | $\ldots$ |
| United Kingdom |  |  |  |  |  |  |  |
| Total | 1,162 | 961 | 688 |  |  | 383 |  |

*Preliminary.

Table 7.2.4 REDFISH. Nominal catches (tonnes) by countries, in Subarea XII 1996-2002, as officially reported to ICES.

| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Estonia | 7,092 | 3,720 | 3,968 | 2,108 | 4,000 | - | - |
| Faroe Islands | 3,127 | 3,822 | 1,793 | 528 |  |  |  |
| France | - | - | 3* | -* | + | 1 | + |
| Germany | 4,391 | 8,866 | 9,746 | 8,204 | 1,128 | 3,833 | 3,032 |
| Greenland | 3,537 | , | 1,180* | 1,188* | 124* | 740* |  |
| Iceland | 3,613 | 3,856 | 1,311 | 5,072 | 3,121 | 11,679 | - |
| Latvia | 1,084 |  | - | - | - | - | 1,144 |
| Norway | 1,013 | 31 | 602 | 2,040 | 2,158* | 878* | 1,094 |
| Poland | - | 662 | - | - | - | - | 1 |
| Portugal | - | - | - | - | - | 387 | -1 |
| Russia | 606 | - | 89 | 7,698 | 9,243 | 4,509 | 6,0382 |
| Spain | 410 | 1,155 | 2,231 | 1,723 | 576 | 1,332 |  |
| UK (E/W/NI) | 33 | - | + | 187 | - | - | $\ldots$ |
| UK (Scotland) | 13 | - | - | 1 | + | - |  |
| United Kingdom |  |  |  |  |  |  | 4 |
| Total | 24,919 | 22,112 | 20,923 | 28,749 |  |  |  |

*Preliminary. ${ }^{1}$ Included in XIV. ${ }^{2}$ See footnote 3 in XIV.

Table 7.2.5
REDFISH. Nominal catches (tonnes) by countries, in Subarea XIV 1996-2002, as officially reported to ICES.

| Country | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | $2002^{*}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Estonia | - | - | - | - | 3,811 | 599 | - |
| Faroe Islands | 298 | 123 | 47 | 2 |  |  |  |
| Germany | 16,996 | 11,610 | 9,709 | 8,935 | 7,840 | 6,758 | 9,576 |
| Greenland | 2,699 | 193 | $296^{*}$ | $3,152^{*}$ | $3,545^{*}$ | $2,587^{*}$ |  |
| Iceland | 49,381 | 33,820 | 6,441 | 23,7701 | 17,999 | 31,786 | 44,430 |
| Norway | 6,453 | 3,187 | 525 | 3,253 | $3,803^{*}$ | $4,258^{*}$ | 4,215 |
| Poland | - | 114 | - | - | - | - | - |
| Portugal | 2,379 | 3,674 | 4,133 | 4,302 | 4,154 | 2,116 | 3,0902 |
| Russia | 45,142 | 36,930 | 25,748 | 16,652 | 14,851 | 23,851 | 25,5423 |
| Spain | 3,897 | 7,552 | 4,660 | 4,175 | 2,657 | 4,982 |  |
| UK (E/W/NI) | 247 | 28 | 43 | 68 | 45 | 179 | $\ldots$ |
| UK (Scotland) | 6 | - | - | - | - | - | $\ldots$ |
| United Kingdom |  |  |  |  |  |  | 33 |
| Total | 127,498 | 97,231 | 51,602 | 64,309 |  |  |  |

${ }^{*}$ Preliminary. ${ }^{1}$ Note Excluding 58 t reported as area unknown. ${ }^{2}$ Reported as V/XII/XIV 3,060 t and 30 t as V/XIV/GRN. ${ }^{3}$ The catch of Atlantic redfishes total of 31,580 tons by ICES subareas XII and XIV, includes catches in NAFO $1 F$ of 4,820 tons.


Figure 7.1.1
Possible relationship between different stocks and species of $S$. marinus and $S$. mentella in the Irminger Sea and adjacent waters.


Figure 7.3.1 Survey abundance indices of Sebastes spp. ( $<17 \mathrm{~cm}$ ) from the German and Icelandic groundfish surveys conducted on the continental shelves of East and West Greenland and Iceland 1985-2001.


Figure 7.5.1 Fishing areas and total catch of the pelagic redfish (S. mentella) by month in 2002, derived from catch statistics provided by Germany, Norway, Iceland, and Greenland. The scale for the catch is in tonnes per squared nautical mile. Total catch for each period is also given.


Figure 7.5.2 Fishing areas and total catch of the pelagic redfish (S. mentella) in the Irmenger Sea and adjacent waters 1995-2002. Data are from Germany (1995-2002), Norway (1995-2002) Greenland (19992002), Russia (1997-2001), Faroese (1995-2001), and Iceland (1995-2002). The scale given is tonnes per square nautical mile.


Figure 7.5.3
Distribution of the Spanish fleet fishing for oceanic redfish (S. mentella) in 2000-2001, divided by Divisions and quarter.


Figure 7.5.4 Position of Russian fleet in the Irminger Sea, divided by month in 2002.




Figure. 7.5.5 Length distribution of the oceanic redfish fishery in ICES Div. XII, XIV and in NAFO Div. 1F by year from 2000-2002. Date from Spain (2000 and 2001) and Russia (2002). The proportion of males is also given.
S.marinus in ICES Divisions V and XIV have been considered as one management unit. Caches in VI have traditionally been included in this report and the group continues to do so.

### 8.1 Landings and trends in the fisheries

Since the early 1980's total catches decreased by more than $70 \%$, from about 130000 t in 1982 to 37000 t in 2001 but increased again in 2002 due to increased catches in Division Va (Table 8.1.1).

In subdivision Va, catches of S. marinus declined from about 63000 t in 1990 to a low of 34000 t in 1996. Since then catches have varied between 35000 and 49000 t , with the lowest catch in 2001 and the highest in 2002. About 90-95\% of the total S. marinus catches in Area Va have in recent years been taken by bottom trawlers (both fresh fish and freezer trawlers; length $48-65 \mathrm{~m}$ ) targeting redfish. The remainder is taken partly as by-catch in gillnet and longline fishery. In 2002, as in previous years, most of the catches were taken along the shelf W, SW, and SE of Iceland, mostly between $12^{\circ} \mathrm{W}$ and $27^{\circ} \mathrm{W}$ (Figure 8.1.1). Although there are no direct measurements available, it is assumed that there are not significant discard of S.marinus in the fishery due to area closures of the most important nursery grounds.

In subdivision Vb , catches have dropped continuously since 1985, from 9000 t to 1500 t in 1999 and has remained at that level since then (Table 8.1.1). Most of the $S$. marinus catches in Vb have been taken by pair trawlers and single trawlers ( $<1000 \mathrm{HP}$ ).

The catches in Division VI increased since 1978, reaching almost 600 t in 1987, followed by a decline to 1992 and have since increased to about 800 t (Table 8.1.1) but decreased again to about 400 t in 2002.

In Division XIV catches have been more variable than in the other areas. Since the highest catch on record ( 31000 t ) in 1982, a rapid decrease was observed to about 2000 t in 1985. During the next 10 years catches varied between 600 and 4200 t . Since 1995 almost no directed fishery for S. marinus occurred and the catches have been 150 t or less. Some by-catch is reported from the shrimp fishery in the area.

### 8.1.1 Biological data form the fishery

The length distributions in the Icelandic landings in 1989-2002 along with measurements from the commercial trawler fleet are shown in Figure 8.1.2. Comparing the length distributions between the catch and landings there are no indications of discard. The numbers of measured fish by statistical square are given in Figure 8.1.3.

Length distribution from the Faroes catches for 2001 and 2002 are shown on Figure 8.1.4. No length data from the catches have been available for last years in Divisions XIV and VI.

The following text-table shows the fishery-related sampling by gear type and Divisions:

| Area | Nation | Gear | Landings | Samples | Fish measured |
| ---: | :--- | :--- | :--- | :--- | :--- |
| Va | Iceland | Bottom trawl | 48,592 | 341 | 62,677 |
| Va | Germany/UK | Bottom trawl | 12 | 0 |  |
| Va | Faroe | Line/hooks | 76 | 0 |  |
| Vb | Faroe | Bottom trawl+gillnets | 1,559 | 29 | 917 |
| XIV | Germany | Bottom trawl | $<150$ | 0 |  |
| VI | UK | Bottom trawl | 392 | 0 |  |

Catch-at-age data from the Icelandic fishery shows that the 1985 year class has dominated the catches from 1995-2001 (Figure 8.2.4 and Table 8.1.2), and in 2002 that year class contributed to $25 \%$ of the total catch in Va. The 1990 year class is also strong and that year class contributed with more than $30 \%$ to the total weight in the catch in 2002 . The average Z , estimated from this 8 -year series of catch-at-age data (Figure 8.1.6) is 0.20 for age groups $15+$, and about 0.18 for age $20+$. This estimation is based on Icelandic age readings, but the ageing can vary between readers. In WD 11 , age reading results are compared between readers in terms of bias and precision. There were significant differences between readers and between methods, mainly for the older fish ( $>20$ years). However, for the medium age range (1120 years), a fairly good agreement was reached between readers. Precision estimates, involving the high longevity of redfish, were relatively good compared to previous age reading comparisons on redfish species.

### 8.2.1 CPUE

CPUE indices for the Icelandic trawl fleet for the period 1985-2002 are estimated from a GLM multiplicative model using summarised data (for each ICES statistical square, vessel, month and year). The model takes into account changes in the Icelandic trawl catches due to vessel, statistical square, month, and year effect. All hauls at depths above 500 m with redfish exceeding $50 \%$ of the total catch, were included in the CPUE estimation (Figure 8.2.1).. A considerable increase in the CPUE was observed in 2001 and is supported by the data from 2002 . The index is now above $85 \%$ of the 1986 value, which is the first year in the series

Unstandardised CPUE from the Fareoese trawler fleet decreased from 1996-1999 by $40 \%$ and remained that low until 2001. The CPUE in 2002 is the lowest value on record (Figure 8.2.2).

### 8.2.2 Survey data

Figure 8.2.3 shows the $S$. marinus abundance index with $95 \%$ confidence intervals using Icelandic Groundfish Survey (IGS) data ( $<400 \mathrm{~m}$ depth). The index is a biomass index of the fishable stock, computed by using a sharp fishable stock ogive (from $34-36 \mathrm{~cm}, \mathrm{~L}_{50}=35 \mathrm{~cm}$ ). In Table 8.2.1 the contribution of each depth stratum to the index is given. The index indicates a decrease in the fishable biomass from 1985-1995, but an increasing trend since then. The lowest index was in 1995 , only about $30 \%$ of the maximum in 1987, but the value in 2003 is $65 \%$ of the highest observed value.

Length distributions from IGS show that the peak (Figure 8.2.4), which has been followed during the last years (first in 1987) now has reached the fishable stock. The increase in the survey index since 1995 therefore reflects the recruitment of a relatively strong year classes (1985 year class and the 1990 year class). This is confirmed by age readings (Figure 8.1.5).

In Division Vb, CPUE of S. marinus were available from the Faroes groundfish survey 1994-2003 (Figure 8.2.5). After an increase in the period from 1995-1998 there is a decrease in 1999-2000. The results also indicate a high variability in the series, and the values are based on average of 43 hauls each year (20-61) hauls.

The new Faroes summer survey (see Section 2) that has been conducted since 1996 shows a constant decreasing trend throughout the series. The index in 2000-2002 is only about $1 / 3$ of the CPUE in 1995 and about the same level as in 1999 and 1998 (Figure 8.2.6).

For the period 1982-2002, abundance and biomass indices from the German groundfish survey for $S$. marinus $>17 \mathrm{~cm}$ are illustrated in Figures 8.2 .7 and 8.2.8. From 1986-1995, an almost continuous reduction in survey biomass has occurred. After a severe depletion of the $S$. marinus stock on the traditional fishing grounds around East Greenland in the early 1990 's, the survey estimates show a significant increase in abundance in 2002 . This increase indicate a possible recovery, although the values are very low compared with the period before 1990. The length frequencies from the German groundfish survey are illustrated in Figures 8.2.9-8.2.10, along with the length distributions in the IGS. The adults seem to remain almost depleted in East Greenland waters.

### 8.2.3 Assessment by use of BORMICON model

Since 1999 the working group has discussed an alternative model (BORMICON (BOReal MIgration and CONsumption model) that has been applied to this stock. The model using S.marinus as an example is described in SCI. MAR,2002 (67 (Suppl. 1): 301-314).

The BORMICON model was run using the same settings as last year's base case. The simulation period is from 1970 to 2003. Two time steps are used each year. Fixed selection pattern is used prior to 1998, but thereafter it is estimated separately for each year. The estimated value of $\mathrm{L}_{50}$ is shown in Table 8.2.4. Results from the runs are shown in Figures 8.2.11-8.2.15 and comparison with last year's results in Figure 8.2.16. As may be seen the stock estimate this year is relatively similar, although a little higher, the difference probably driven by the high survey index in 2003. Survey indices have varied much but have in general been increasing since 1993-1995 and reach its highest value in 2003. The survey index used here (Table 8.2.4) is total biomass index and differs therefore form the index of fishable biomass shown in Figure 8.2.3. Indices have been attended with relatively high CV's (Table 8.2.4). The CV of the survey index in 2003 is similar as the average CV in last 10 years.

Natural mortality is set to 0.15 for the youngest, decreasing gradually to 0.05 for age 5 and older. Alternatives with other values on natural moralities ( $\mathrm{M}=0.1$ for age $5+$ ) were tested in 2002. They gave a worse fit and are therefore not tried again this year. The ages used are 1 to 30 years. The oldest age is treated as a plus group. Recruitment was at age 1. Prior to 1989 length at recruitment was 7.1 cm , but 8.1 cm in later years. This was supposed to reflect the length of the 1985 and 1990 year classes in the groundfish survey.

Figure 8.2.15. shows residuals from the model fit to the survey data, demonstrating large positive residuals in some years, most notably 1993,1999 and in 2003.

The IGS in 2003 does not indicate any improvement in redfish recruitment, which has been bad since 1990-1991. The estimated average year class size in 1992-2001 is estimated 80 million (at age 0 ). Maximum yield-per-recruit is 250 g , so this recruitment can only sustain an annual catch of 20000 tonnes. According to the predictions here, the stock is going to be stable for the next few years with an annual catch of 35-40 000 tonnes. This value might have to be reduced every year, though, when no sign of good recruitment is seen. From the above-mentioned runs, it is clear that if the groundfish survey is to be accepted as a measure of recruitment, no new large year class will recruit to the fishable stock within the next 10 years.

In 2001 the model was also run with the total S.marinus catch in ICES Divisions XIV, Va, and Vb. This addition increased the estimated stock size as the catch increased. Nevertheless, the proportion of the catch taken in Division Va has been relatively stable since 1985 , with about $85-90 \%$ taken in Va. As the tuning data are identical, similar trends in the stock size are to be expected in recent years, with about $10 \%$ higher biomass in 2000 than when using only the data from Va.

The main indicator for recruitment is the groundfish survey, which does not indicate any strong year class after the 1990/1991 year class. Simulations were used to determine the value of $\mathbf{F}_{\text {max }}$. A year class was started in 1970 and caught using fixed fishing mortality and the estimated selection pattern. The total yield from the year class was then calculated. $\mathbf{F}_{\text {max }}$ was calculated at 0.165 using 40 years simulations, and $\mathbf{F}_{0.1}$ was estimated to be 0.09 . Here, F is not fishing mortality, but close to it when small time steps are used, or moralities are small. It is also the mortality of a fish where the selection is 1 .

Different catch options were tested in the future simulations for a fixed catch. As may be seen in Figure 8.2.13-8.2.14, the catchable biomass will increase until 2005, using fixed catch after the year 2002 for all catch options below 40000 t . The total biomass will at the end of the period be lower than it is now for catches exceeding about 35000 t annually.

### 8.2.4 State of the stock

All available survey information and CPUE data from Division Va show that the $S$. marinus stock decreased considerably from 1985 to the lowest recorded biomass in 1995. An improvement in the fishable biomass has, however, been seen in the most recent years due to improved recruitment. During the last few years, the 1985 year class has contributed significantly to the fishable stock, and the 1990 year class has also contributed significantly to the fishable biomass in the last 3 years. It is expected that those year classes will dominate the catches in the next few years. However, there is no indication of new, strong year classes after the 1990 year class. In Vb , survey indices as well as CPUE from the fleet do not indicate improved situation in the area and adult fish in Subarea XIV has nearly been exhausted in the most recent years but there are signs of improved recruitment (Figure 8.2.10).In summary, the Icelandic groundfish survey, as well as the CPUE series, seem to indicate a considerable decline in the fishable biomass of $S$. marinus during the period from 1986 to 1994. The stock has increased, and is now inside defined safe biological limits $\left(\mathrm{U}_{\mathrm{pa}}\right)$. A large proportion of the catches in recent years is caught from only two year classes. The fishable stock situation remains bad for Division XIV and Vb.

### 8.2.5 Catch projections and management considerations

The Icelandic groundfish survey indices ( U ) may be assumed to be related to overall biomass (B) by a simple linear relationship $(\mathrm{U}=\mathrm{kB})$. If catches are assumed to be proportional to stock size and effort $(\mathrm{Y}=\mathrm{cEB})$, then it follows that catch over survey index is proportional to effort $(\mathrm{Y} / \mathrm{U}=\mathrm{aE}$, see Table 8.2.2) and this allows a one-year prediction of catch, assuming a status-quo effort level. Although calculated confidence limits in the groundfish survey are quite low, year-to-year variation in catchability/availability will affect the results drastically while using only the last observation value as a basis for extrapolation of catches in the coming year, based on a constant effort. By using a running average over a few years ( 3 as a minimum), one would reduce the variation in the catch prediction, based on the above assumptions.

By assuming same effort in 2004 as in 2002 (see Section 8.2.1) the predicted catch in Va will be around 48000 tonnes, using the formula, Catch $_{2004}=$ Average Survey index $x_{2001-2003} *$ Effort $_{2002}$. By applying the same method for Vb, using commercial CPUE data (both series combined) instead of survey index, a predicted catch in Vb would be around 1900 tonnes by assuming the same effort in 2004 as it was in 2002.

The ACFM formulation for advice in last 2 years was to reduce the effort by $25 \%$ based on the approach given above. That corresponded to 31000 tonnes in Division V for 2003. By applying the same method for 2004, the catches would be about 36000 in Division Va and around 1400 t in Vb.

Based on the BORMICON model, a decrease in the fishable biomass is expected for all catch options above about 40000 t . This is due to the poor recruitment after the 1990 year class. The estimated average year class since 1992 is about 80 millions (at age 0 ) and maximum yield-per-recruit is about 250 gr . A large proportion of the catch is from two year classes from 1985 and 1990. Therefore, after these two strong year classes have passed the fishery, one cannot expect higher yield than about 20000 t from the year classes that come into the fishable stock in the next years. Based on the results, a TAC below 40000 t in the next 5 years would provide a fishable stock size above current biomass level, at the end of that period. The approximate F from the model would increase after 2004 but fishing mortality would be above $\mathbf{F}_{\max }$. Catches corresponding to $\mathbf{F}_{\max }$ in next 5 years would be around 35000 tonnes.

In order to rebuild the stock further in the near future, effort should be kept low.

### 8.3 Biological reference points

S. marinus is mainly caught in Division Va, and the relative state of the stock can be assessed through survey index series from that Division. ACFM accepted the proposal of the working group of defining reference points in terms of current state with respect to $\mathrm{U}_{\mathrm{lim}}=\mathrm{U}_{\text {max }} / 5$ and $\mathrm{U}_{\mathrm{pa}}=60 \%$ of $\mathrm{U}_{\text {max }}$. $\mathrm{U}_{\mathrm{pa}}$ corresponds to the fishable biomass associated with the last strong year class. Based on survey data, the highest recorded biomass was reached in 1987. Based on these definitions, the stock has been close to Upa during the last years. The survey index series is only available back to 1985 .

The group discussed other alternatives to define reference points for this species, such as F or B points based on the BORMICON model. Although the BORMICON model provides candidates for reference points, they are difficult to use as the most important input data to the model is short, compared to the live span of the species and no year class has been followed throughout the fishery. The strong year classes that were observed in the first surveys are still the most dominant year classes in the fishery. Therefore the group considered it not appropriate at this time to change the biological reference points that have been used.

### 8.4 Comment on the assessment

The BORMICON model used for this stock extracts data directly from the databases at the Marine Research Institute in Reykjavik and intermediate input/output files are not included. This severely reduces the ability of the WG to review data and model results. Relevant intermediate input/output files and model diagnostic should be made available for the next assessment.

There are only available data on nursery grounds of S.marinus in Icelandic and Greenlandic waters but no nursery grounds are known in the Faroe Islands area. In Icelandic waters, nursery areas are found mostly West and North of Iceland at depths between 50 and approximately 350 m , but also in the South and East (ICES C.M. 1983/G:3; Einarsson, 1960; Magnússon and Magnússon 1975; Pálsson et al. 1997). As the length (age) increases, migration of young S.marinus is anticlockwise from the North coast to the West coast and further to the Southeast fishing areas and to Faroese fishing grounds in Vb . The largest specimens are found in Subdivision Vb and therefore the year classes from 1985 and 1990 might still not have entered into that area. This might explain the inconsistency between different indicators on the status of the stock.

Table 8.1.1 S. marinus. Landings (in tonnes) by area used by the Working Group.

|  | Area |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Va | Vb | VI | XII | XIV | Total |
| 1978 | 31,300 | 2,039 | 313 | 0 | 15,477 | 49,129 |
| 1979 | 56,616 | 4,805 | 6 | 0 | 15,787 | 77,214 |
| 1980 | 62,052 | 4,920 | 2 | 0 | 22,203 | 89,177 |
| 1981 | 75,828 | 2,538 | 3 | 0 | 23,608 | 101,977 |
| 1982 | 97,899 | 1,810 | 28 | 0 | 30,692 | 130,429 |
| 1983 | 87,412 | 3,394 | 60 | 0 | 15,636 | 106,502 |
| 1984 | 84,766 | 6,228 | 86 | 0 | 5,040 | 96,120 |
| 1985 | 67,312 | 9,194 | 245 | 0 | 2,117 | 78,868 |
| 1986 | 67,772 | 6,300 | 288 | 0 | 2,988 | 77,348 |
| 1987 | 69,212 | 6,143 | 576 | 0 | 1,196 | 77,127 |
| 1988 | 80,472 | 5,020 | 533 | 0 | 3,964 | 89,989 |
| 1989 | 51,852 | 4,140 | 373 | 0 | 685 | 57,050 |
| 1990 | 63,156 | 2,407 | 382 | 0 | 687 | 66,632 |
| 1991 | 49,677 | 2,140 | 292 | 0 | 4,255 | 56,364 |
| 1992 | 51,464 | 3,460 | 40 | 0 | 746 | 55,710 |
| 1993 | 45,890 | 2,621 | 101 | 0 | 1,738 | 50,350 |
| 1994 | 38,669 | 2,274 | 129 | 0 | 1,443 | 42,515 |
| 1995 | 41,516 | 2,581 | 606 | 0 | 62 | 44,765 |
| 1996 | 33,558 | 2,316 | 664 | 0 | 59 | 36,597 |
| 1997 | 36,342 | 2,839 | 542 | 0 | 37 | 39,761 |
| 1998 | 36,771 | 2,565 | 379 | 0 | 109 | 39,825 |
| 1999 | 39,824 | 1,436 | 773 | 0 | 7 | 42,040 |
| 2000 | 41,187 | 1,498 | 776 | 0 | 89 | 43,550 |
| 2001 | 34,895 | 1,489 | 535 | 0 | 93 | 37,012 |
| 2002 | 48,648 | 1,559 | 392 | 0 | 189 | 50,788 |
|  |  |  |  |  |  |  |

Table 8.1.2 S. marinus. Catches in Va in weight (tonnes) by age.

| Year/ <br> Age | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | ---: | ---: |
| 7 | 59 | 0 | 33 | 24 | 0 | 0 | 125 | 0 |
| 8 | 366 | 354 | 229 | 285 | 367 | 118 | 140 | 631 |
| 9 | 1572 | 808 | 483 | 598 | 1492 | 595 | 396 | 653 |
| 10 | 9312 | 3622 | 1039 | 1213 | 1244 | 3977 | 1625 | 484 |
| 11 | 2698 | 8943 | 2704 | 1134 | 1820 | 1894 | 7757 | 2661 |
| 12 | 1314 | 2072 | 11563 | 3257 | 2651 | 2524 | 1804 | 12744 |
| 13 | 3548 | 1300 | 2820 | 12548 | 2330 | 1610 | 1978 | 2188 |
| 14 | 5684 | 1459 | 1366 | 2086 | 15703 | 2292 | 1249 | 2160 |
| 15 | 6000 | 4398 | 3123 | 2039 | 1171 | 14272 | 836 | 1524 |
| 16 | 1743 | 5641 | 3621 | 2411 | 1235 | 1778 | 11649 | 2379 |
| 17 | 859 | 921 | 3024 | 3410 | 1884 | 1234 | 521 | 14412 |
| 18 | 371 | 388 | 896 | 2048 | 2769 | 1843 | 784 | 1646 |
| 19 | 1148 | 268 | 644 | 1015 | 2317 | 2379 | 1064 | 1325 |
| 20 | 1158 | 337 | 960 | 726 | 1219 | 2201 | 1794 | 835 |
| 21 | 511 | 1210 | 448 | 521 | 487 | 571 | 966 | 972 |
| 22 | 684 | 1033 | 544 | 390 | 231 | 619 | 418 | 833 |
| 23 | 1447 | 803 | 691 | 425 | 347 | 226 | 435 | 675 |
| 24 | 673 | 0 | 595 | 662 | 226 | 124 | 168 | 88 |
| 25 | 773 | 0 | 753 | 516 | 948 | 585 | 130 | 541 |
| 26 | 370 | 0 | 271 | 400 | 281 | 503 | 125 | 88 |
| 27 | 354 | 0 | 140 | 425 | 587 | 248 | 291 | 222 |
| 28 | 736 | 0 | 208 | 359 | 175 | 493 | 204 | 616 |
| 29 | 0 | 0 | 155 | 54 | 107 | 471 | 153 | 553 |
| 30 | 134 | 0 | 31 | 226 | 234 | 451 | 373 | 350 |

Table 8.2.1 Index on fishable stock of $S$. marinus in the Icelandic groundfish survey by depth.

| Depth interv |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year < 100m |  |  | 200-400m | 400-500m | Total 0-400m | Total |
| 1985 | 7 | 91 | 140 | 24 | 237 | 261 |
| 1986 | 2 | 86 | 180 | 12 | 268 | 280 |
| 1987 | 2 | 124 | 150 | 10 | 276 | 286 |
| 1988 | 1 | 95 | 110 | 4 | 206 | 210 |
| 1989 | 1 | 101 | 118 | 11 | 220 | 231 |
| 1990 | 2 | 68 | 81 | 22 | 151 | 173 |
| 1991 | 2 | 76 | 53 | 8 | 130 | 139 |
| 1992 | 1 | 62 | 59 | 9 | 122 | 132 |
| 1993 | 1 | 48 | 50 | 17 | 98 | 115 |
| 1994 | 1 | 58 | 51 | 1 | 110 | 111 |
| 1995 | 0 | 36 | 45 | 11 | 81 | 92 |
| 1996 | 1 | 44 | 76 | 21 | 122 | 143 |
| 1997 | 1 | 60 | 71 | 34 | 133 | 166 |
| 1998 | 2 | 57 | 71 | 3 | 130 | 132 |
| 1999 | 1 | 56 | 107 | 44 | 164 | 208 |
| 2000 | 2 | 47 | 69 | 8 | 117 | 125 |
| 2001 | 2 | 33 | 67 | 6 | 101 | 107 |
| 2002 | 2 | 64 | 74 | 11 | 140 | 151 |
| 2003 | 9 | 60 | 107 | 29 | 176 | 205 |

Table 8.2.2 S. marinus. Results from the Icelandic groundfish survey in Va, total catch in Va and effort towards $S$. marinus.

| Year | Survey <br> index | Catch <br> (Va) | Effort |
| :--- | :--- | :--- | :--- |
| 1985 | 1000 | 67,312 | 67 |
| 1986 | 1131 | 67,772 | 60 |
| 1987 | 1165 | 69,212 | 60 |
| 1988 | 869 | 80,472 | 93 |
| 1989 | 928 | 51,852 | 56 |
| 1990 | 637 | 63,156 | 99 |
| 1991 | 549 | 49,677 | 91 |
| 1992 | 515 | 51,464 | 100 |
| 1993 | 414 | 45,890 | 111 |
| 1994 | 464 | 38,669 | 84 |
| 1995 | 342 | 41,516 | 122 |
| 1996 | 511 | 33,558 | 66 |
| 1997 | 561 | 36,342 | 65 |
| 1998 | 549 | 36,771 | 67 |
| 1999 | 692 | 39,824 | 58 |
| 2000 | 494 | 41,110 | 83 |
| 2001 | 426 | 34,986 | 82 |
| 2002 | 590 | 48,648 | 82 |
| 2003 | 741 |  |  |

Table 8.2.3 Results of the BORMICON model. BASE CASE, estimated value of $\mathrm{L}_{50}$.

| Year | $<1998$ | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $\mathrm{~L}_{50}$ | 34.18 | 34.81 | 34.55 | 34.10 | 33.67 | 33.71 |

Table 8.2.4 Index of total biomass of S.marinus from the groundfish survey in March and CV in the estimate.

| Year | index | CV |
| :--- | :--- | :--- |
| 1985 | 332 | 0.094 |
| 1986 | 373 | 0.134 |
| 1987 | 349 | 0.114 |
| 1988 | 283 | 0.103 |
| 1989 | 337 | 0.152 |
| 1990 | 311 | 0.313 |
| 1991 | 203 | 0.105 |
| 1992 | 173 | 0.093 |
| 1993 | 203 | 0.143 |
| 1994 | 189 | 0.125 |
| 1995 | 163 | 0.138 |
| 1996 | 228 | 0.211 |
| 1997 | 279 | 0.311 |
| 1998 | 228 | 0.159 |
| 1999 | 377 | 0.206 |
| 2000 | 261 | 0.203 |
| 2001 | 221 | 0.154 |
| 2002 | 255 | 0.122 |
| 2003 | 402 | 0.191 |



Figure 8.1.1 Distribution of S.marinus catches in Division Va from 2000-2002.


Figure 8.1.2
S. marinus. Length distribution from Icelandic landings and from samples taken at sea from the trawler fleet 1989-2002.
yfirlit_synasofnun_teg_5_year_2002.ps


Figure 8.1.3 Number of measured $S$. marinus in 2002 by statistical square.


Figure 8.1.4 S. marinus. Length distribution from Faroese catches in 2001-2002.









Figure 8.1.5 S. marinus. Catch in number by age in ICES Subdivision Va 1995-2002.


Figure 8.1.6 S. marinus. Catch curve based on the catch data in ICES Division Va 1995-2002.


Figure 8.2.1 CPUE in $S$. marinus from Icelandic trawlers, both based on results from the GLM model 19852002 with $95 \%$ CV where the S. marinus catch composed $50 \%$ or more of the total catch in each haul.


Figure 8.2.2 CPUE from the Faroese pair-trawlers in ICES Division Vb 1985-2002.


Figure 8.2.3 Index on fishable stock of $S$. marinus from Icelandic groundfish survey and $95 \%$ confidence intervals. The index is based on all strata at depths from 0-400 m.


Figure 8.2.4 Length distribution of S. marinus in the Icelandic groundfish survey 1985-2003.


Figure 8.2.5 CPUE of S.marinus in the Faroes groundfish survey 1994-2003.


Figure 8.2.6
CPUE of $S$. marinus in the Faroes summer survey in Division Vb1 from 1996-2002.


Figure 8.2.7 S. marinus ( $\geq 17 \mathrm{~cm}$ ). Survey abundance indices for East, West Greenland and Iceland 19852002.


Figure 8.2.8
S. marinus ( $\geq 17 \mathrm{~cm}$ ). Survey biomass indices for East and West Greenland and Iceland, 1985-2002.


Figure 8.2.9
S. marinus (>17cm). Length frequencies for East Greenland, West Greenland, and Iceland, 19851994.


Figure 8.2.10 S. marinus ( $>17 \mathrm{~cm}$ ). Length frequencies for East Greenland, West Greenland, and Iceland, 19952002.

Estimated selection pattern



Effects of catch on growth


Recruitment at age 0 in million fishes


Figure 8.2.11
Results from the BORMICON model-BASE CASE, using catch data from ICES Division Va only. a) Estimated selection pattern of the commercial fleet and the survey, b) Mean length (the Figure also demonstrates the effect of catch on length-at-age), c) Yield-per-recruit, and d) Estimated recruitment at age 0 .

Total biomass in 1000 tonnes


Fishing mortality of age 20


Catch in 1000 tonnes/year



Figure 8.2.12
Results from the BASE CASE run, using only catch data from ICES Division Va. The Figures show the development of biomass and $F$, using different catch options ( $0-60000 \mathrm{t}$ ) after 2003.


Figure 8.2.13 Results from the BASE CASE run, using only catch data from ICES Division Va. The Figures show the development of biomass and F, using different effort after 2003.


Figure 8.2.14 Results from the BASE CASE run, using only catch data from ICES Division Va. The Figure shows comparison of observed and modelled survey biomass (total biomass).


Figure 8.2.15
Results from the BASE CASE run, using only catch data from ICES Division Va. Residuals from fit to survey data $\log$ (Isur/Imod). The shaded circles show positive residuals (survey results exceed model prediction).


Figure 8.2.16 Comparison of catchable biomass using the data obtained now and last year, for same settings. Results are obtained using only the catch history from ICES Division Va.

Deep-sea $S$. mentella on the continental shelves and slopes around the Faroe Islands, Iceland, and East Greenland is treated as one stock unit and separated from the stock fished in the Irminger Sea (oceanic S. mentella, see Chapter 10). It is believed to have a common area of larval extrusion southwest of Iceland, a drift of the pelagic fry towards the nursery areas on relatively shallow waters off East Greenland, and feeding and copulation areas on the shelves and banks around the Faroe Islands, Iceland, and East-Greenland. The main fishing grounds are in Icelandic waters.

### 9.1 Landings and Trends in the Fisheries

The total annual landings of deep-sea $S$. mentella from Divisions Va and Vb, and Subareas VI and XIV varied between 20,000 and 84,000 t in 1978-1994 (Table 9.1.1). Since 1994, landings gradually decreased and in 2001 catches were $23,000 \mathrm{t}$, which was the lowest recorded catch since 1979. Catches in 2002 were similar to 2001.

In Division Va, annual landings gradually decreased from a record high of $57,000 \mathrm{t}$ in 1994 to $17,000 \mathrm{t}$ and $19,000 \mathrm{t}$ in 2001 and 2002 respectively. For the past three years most of the catch were taken by bottom trawlers (the pelagic trawl fishery was in 2002 negligible) along the shelf west, southwest, and southeast of Iceland (Figure 9.1.1), east of so called redfish line (Figure 9.1.2). The catches in the third and fourth quarter of the year decreased considerable in 2001 compared with the years before and continued to decrease in 2002 (Figure 9.1.3). The reason for this decrease seems to be associated with decreased effort at that time of year. Some fishermen believe that this is because of much less deepsea $S$. mentella in the traditional fishing areas west and southwest of Iceland. Length distribution of deep-sea $S$. mentella from the bottom trawl fishery shows an increase in the number of small fish in the catch in recent years (Figure 9.1.4). A peak of about 32 cm in 1994 can be followed by approximately 1 cm annual growth in 1996-2002.

In Division Vb , landings of deep-sea $S$. mentella were $2,700 \mathrm{t}$ in 2002, which is a decrease from 4,200 t in 2001. Maximum annual landings were $15,000 \mathrm{t}$ in 1986. Length distribution from the catch in 2001 and 2002 indicates that the fish caught are larger than 40 cm (Figure 9.1.5).

In Subarea VI, the annual landings varied between 200 t and 1,100 t 1978-2000 (Table 9.1.1). Catch statistics for 2001 and 2002 indicate that the redfish fishery in VI has ceased to a very low level. About 20 t were caught in 2002.

In Subarea XIV, the annual deep-sea S. mentella catch has decreased drastically. In 1980-1994, catches varied between 2,000 and 19,000 with the lowest catch in 1989 and the highest in 1994. In the following three years, the annual landings were less than $1,000 \mathrm{t}$ and the redfish was mainly caught as by-catch in the shrimp fishery. In 1998, Germany started a directed fishery for redfish with annual landings around $1,000 \mathrm{t}$ 1998-2001, but landings increased to $1,900 \mathrm{t}$ in 2002. Samples taken from the German fleet indicates that substantial quantities of the fish caught, especially in 2002, are juveniles, i.e. fish less than 30 cm (Figure 9.1.6)

Below is a text table showing the 2002 biological sampling from the catch and landings of deep-sea $S$. mentella from the continental shelf divided by Division and gear type.

| Area | Gear | Landings | Nos. samples | Nos. fish measured |
| :--- | :--- | :--- | :--- | :--- |
| Va | Pelagic trawl | 44 | 0 | 0 |
| Va | Bottom trawl | 19,100 | 172 | 30,197 |
| Vb | Bottom trawl | 2,674 | 38 | 4,424 |
| XIVb | Bottom trawl | 1,903 | 1 | $?$ |

### 9.2 Assessment

### 9.2.1 Trends in CPUE and survey indices

Data used to estimate CPUE for deep-sea S. mentella in Division Va 1986-2002 were obtained from log-books of the Icelandic trawl fleet. Only bottom trawl tows taken below 500 m depth were used and where $S$. mentella composed at least $50 \%$ of the total catch in each tow. Indices of CPUE were estimated from this data set using a GLM multiplicative model. This model takes into account changes in vessels over time as well as difference in vessel size, area (ICES statistical square), and month and year effects.

From 1986 to 1989 CPUE in Division Va was relatively stable, but gradually decreased from 1989 to a record low in 1994 (Figure 9.2.1). From 1995 to 2000, CPUE slightly increased annually, but decreased again in 2001 and 2002. The
fishing effort at the time when the stock was considered in stable condition, i.e. from 1986-1990, was $15,000-40,000$ hours fishing. From 1991 to 1994, the fishing effort increased drastically, but decreased between $10 \%$ and $20 \%$ each year to 2001. ICES recommended $25 \%$ annual reduction in fishing effort during the same time period. Effort increased between 2001 and 2002 by about $10 \%$.

CPUE indices in Division Vb for deep-sea $S$. mentella and obtained from the Faroese CUBA trawlers decreased from $500 \mathrm{~kg} /$ hour in 1991 to of $300 \mathrm{~kg} /$ hour in 1993 (Figure 9.2.2). CPUE was at this level until 2002 when it increased to $370 \mathrm{~kg} / \mathrm{hour}$. Fishing effort during this period decreased between 2001 and 2002 as recommended by the NWWG in 2002. The summer survey 1996-2002 in Division Vb shows nearly a continuous decrease in the catch rate or from less than $10 \mathrm{~kg} / \mathrm{hour}$ to about $3 \mathrm{~kg} / \mathrm{h}$ (except in 1999 when the catches were over $10 \mathrm{~kg} / \mathrm{h}$ ) (Figure 9.2.3).

CPUE data from Division XIV were only available from 1998 when directed fishery for $S$. mentella by Germany started along the continental slope of East Greenland. Fishing effort was similar in the first three years or around 2,200 hours fishing, decreased to 1,000 hours in 20001, and increased again in 2002 to 1,500 hours (Table 9.2.1). At the same time, CPUE decreased between 1998 and 1999, but has since then increased annually.

Surveys conducted on the continental shelf of West and East Greenland and Iceland cover only the distribution of juvenile deep-sea $S$. mentella (recruits). The results indicate that juveniles are most abundant off East Greenland, while negligible part of juveniles are distributed off West Greenland and Iceland (Figure 9.2.4). Figure 9.2.4 shows that the abundance was dominated by a single strong year class recorded for the first time in 1987 at a mean length of 20 cm . Annual growth of this cohort was about 2 cm and fully recruited to the survey gear in 1997 at a length of about 27 cm , when abundance and biomass reached its maximum (total abundance estimated 7 billion individuals and biomass 1.5 million tons). This year class seems to have left the survey area in the following years. In recent years there is an indication of recruiting year classes, but they seem, however, to be significantly less abundant. Abundance (Figure 9.2 .5 ) and biomass (Figure 9.2.6) indices have been gradually increasing since 1999, but are only about 20\% of what they were in 1997.

### 9.2.2 Stock production model

As described in previous report, the group used the ASPIC stock production model (Prager 1992) for the stock, fitted to CPUE indices and catch. It should be noted that he model runs are exploratory and are not used as basis for advice of the stock. The model requires catch data and corresponding fishing effort or CPUE data that are indices of the stock biomass. Corresponding CPUE data were available from the Icelandic trawl fishery 1986-2002 in Division Va and from the Faroese Islands CUBA fishery 1991-2002. The total catch from Divisions V, VI, and XIVb 1986-2002 was used in the model run. Results are given in Table 9.2.2.

MSY was estimated $48,700 \mathrm{t}$ and $\mathbf{B}_{\text {MSY }} 243,200 \mathrm{t}$. Biomass in 2002 is estimated to be over $30 \%$ above $\mathbf{B}_{\text {MSY }}$ and the fishing mortality is estimated to be about $40 \%$ below $\mathbf{F}_{\text {MSY }}$. Observed and estimated CPUE's are given in Table 9.2.2. State of the stock relative to retrospective runs show rather stable situation, but indicates that $\mathrm{B} / \mathbf{B}_{\mathrm{MSY}}$ has been underestimated and $\mathrm{F} / \mathbf{F}_{\mathrm{MSY}}$ overestimated systematically in recent years (Figure 9.2.7). The biomass is slowly increasing from the record low in 1995. Although the retrospective runs show relatively consistency between years, the parameters that are used as a basis for calculating MSY, $\mathbf{F}_{\mathrm{MSY}}$, and $\mathbf{B}_{\mathrm{MSY}}(\mathrm{K}$ and r ) are changing over time (Table 9.2.3). Plot of B-ratio (B/B $\mathbf{B}_{\mathrm{MSY}}$ ) along with biomass trajectory is given in Table 9.2.2. Fishing at $\mathbf{F}_{\mathrm{pa}}\left(2 / 3 \mathbf{F}_{\mathrm{MSY}} \sim \mathbf{F}_{\mathrm{pa}}\right)$ will keep the stock well above $\mathbf{B}_{\mathrm{MSY}}$ within the next years. The ASPIC model results shows that catching at $\mathbf{F}_{\mathrm{pa}}$ would result in a total catch around $32,500 \mathrm{t}$ in 2003. Those catches apply to the whole distribution area of $S$. mentella.

The intrinsic rate of population increase, $r$, was estimated 0.40 and it might be too high for such a long-lived and slowgrowing species. However, one should not take the estimates of $r$ from a production model, such as ASPIC, as informative as of the actual intrinsic rate of increase of a population, as it may be poorly estimated. Simple production models are much better at estimating MSY and effort at MSY ( $\mathbf{F}_{\mathrm{MSY}}$ ) than they are at estimating $r$, K, or fishing mortality rate at MSY ( $\mathbf{F}_{\text {MSY }}$ ). This arises because for most populations, the data don't reveal whether MSY can be taken as a larger fraction of a smaller population or vice versa (Michael Prager, personal communication). $r$ is estimated similar to the value obtained for the species and related species in FishBase..

### 9.2.3 State of the stock

All CPUE indices in Division Va show a drastic reduction from highs in the late 1980s, but there is an indication that the stock have started a slow recovery since the middle of 1990s, when CPUE was close to $50 \%$ of the maximum. Fishermen have, however, reported of less $S$. mentella in the fishing areas southwest and west of Iceland in past two years compared to the most recent years. The increase in the fishable biomass in Division Va since 1996 may be related
to recruiting fish, most likely from East Greenland. It is, however, uncertain to what extent the juvenile $S$. mentella observed at East Greenland will recruit to this stock.

In Division Vb development in CPUE resembles that in Division Va, i.e., the CPUE seems to have stabilised at below $50 \%$ of the maximum in the time-series. There was, however, an increase in CPUE in 2002 compared to the year before.

Based on survey results, the fishable stock of deep-sea $S$. mentella on the continental shelf in area XIV is severely depleted (Figure 9.2.5). The strong recruiting cohort(s) observed in 1993-97 emigrated from the area in 1998-2000 and may have recruited the stock on the shelf.

### 9.3 Catch projections

It is possible to estimate catch for deep-sea $S$. mentella that corresponds to different effort. Here, this was done by using the formula:

$$
\text { Catch }_{2004}=\text { Average CPUE index } 2000-2002 * \text { Average Effort } 2000-2002
$$

where effort for each year is calculated as: Effort = Catch/CPUE. This formula was applied to catch statistic and CPUE from Va and Vb , giving a catch of $22,314 \mathrm{t}$ in Va and $4,065 \mathrm{t}$ in Vb (Table 9.2.4). This will correspond to a catch of $26,380 \mathrm{t}$ for the whole stock (Table 9.2.4). The above calculation was a basis for the ACFM advice last year.

### 9.4 Biological reference points

The relative state of the stock can be assessed through CPUE index series (U) from the commercial fishery, which imply a maximum, $U_{\text {max }}$. It has been proposed by ACFM that reference points can be defined in terms of the current state with respect to $U_{\lim }=U_{\max } / 5$ and $U_{p a}=U_{\max } / 2$. Based on these definitions, the stock is considered above $U_{p a}$.

### 9.5 Management considerations

The catch increased between 2001 and 2002 in Division Va, but was considerable lower than the set quota. This could be due to increased effort towards other species, such as $S$. marinus, in relation to depleted stock of $S$. mentella. It should be noted that Icelandic authorities give a joint quota for $S$. marinus and $S$. mentella. The working group recommends that the TAC of demersal redfish stocks ( $S$. marinus and deep-sea $S$. mentella) should be given separately. There is a strong indication that these two species in Va are spatially separated and therefore, separate quotas for these species can be given.

In recent years, ICES has advised that the fishing effort towards $S$. mentella should kept low and both in Division Va and Vb fishing effort has decreased considerable. The management strategy to reduce the effort has resulted in an increase in the fishable biomass since 1995 according the data from the fishery. The WG recommends that effort in should be kept low and no higher than it was in 2000-2002. That corresponds to a catch of 22,314 tin Va in 2004 and $4,065 \mathrm{t}$ in Vb .

The Working Group recommends maximum protection of the juveniles in Division XIV and therefore, no directed fishery towards deep-sea $S$. mentella should be allowed.

Table 9.1.1 Deep-sea $S$. mentella. Nominal catch (tonnes) on the continental shelf and slopes divided by ICES area.

|  | Area |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Va | Vb | VI | XII | XIV | Total |
| 1978 | 3,902 | 7,767 | 18 | 0 | 5,403 | 17,090 |
| 1979 | 7,694 | 7,869 | 819 | 0 | 5,131 | 21,513 |
| 1980 | 10,197 | 5,119 | 1,109 | 0 | 10,406 | 26,831 |
| 1981 | 19,689 | 4,607 | 1,008 | 0 | 19,391 | 44,695 |
| 1982 | 18,492 | 7,631 | 626 | 0 | 12,140 | 38,889 |
| 1983 | 37,115 | 5,990 | 396 | 0 | 15,207 | 58,708 |
| 1984 | 24,493 | 7,704 | 609 | 0 | 9,126 | 41,932 |
| 1985 | 24,768 | 10,560 | 247 | 0 | 9,376 | 44,951 |
| 1986 | 18,898 | 15,176 | 242 | 0 | 12,138 | 46,454 |
| 1987 | 19,293 | 11,395 | 478 | 0 | 6,407 | 37,573 |
| 1988 | 14,290 | 10,488 | 590 | 0 | 6,065 | 31,433 |
| 1989 | 40,269 | 10,928 | 424 | 0 | 2,284 | 53,905 |
| 1990 | 28,429 | 9,330 | 348 | 0 | 6,097 | 44,204 |
| 1991 | 47,651 | 12,897 | 273 | 0 | 7,057 | 67,879 |
| 1992 | 43,414 | 12,533 | 134 | 0 | 7,022 | 63,103 |
| 1993 | 51,221 | 7,801 | 346 | 0 | 14,828 | 74,196 |
| 1994 | 56,720 | 6,899 | 642 | 0 | 19,305 | 83,566 |
| 1995 | 48,708 | 5,670 | 536 | 0 | 819 | 55,733 |
| 1996 | 34,741 | 5,337 | 1,048 | 0 | 730 | 41,856 |
| 1997 | 37,876 | 4,558 | 419 | 0 | 199 | 43,051 |
| 1998 | 33,125 | 4,089 | 298 | 3 | 1,376 | 38,890 |
| 1999 | 28,590 | 5,294 | 243 | 0 | 865 | 34,992 |
| 2000 | 30,696 | 4,841 | 885 | 0 | 986 | 37,408 |
| 2001 | 17,313 | 4,247 | 34 | 0 | 927 | 22,521 |
| 2002 | 19,148 | 2,674 | 19 | 0 | 1,903 | 23,744 |
|  |  |  |  |  |  |  |

Table 9.2.1 S. marinus and deep sea S. mentella nominal catch (tonnes) and effort (hours fished) of the German fleet in ICES Divisions Va, Vb, VI and XIVb by year and quarter 1998-2002.

| Year | Quarter | Total |  |  | Va |  | Vb |  | VI |  |  | XIVb |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Catch (t) | Effort (h) | Catch (t) | Effort (h) | Catch (t) | Effort (h) | Catch (t) |  | Effort (h) | Catch (t) | Effort (h) |
| 1998 | 1 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |
| 1998 | 2 | 2 | 290 | 629 | 0 | 0 | 0 | 0 |  | 0 | 15 | 290 | 614 |
| 1998 | 3 | 3 | 401 | 1202 | 226 | 945 | 0 | 0 |  | 0 | 0 | 175 | 260 |
| 1998 | 4 | 4 | 1001 | 1628 | 58 | 287 | 0 | 0 |  | 0 | 6 | 943 | 1335 |
| 1998 |  |  | 1693 | 3459 | 284 | 1232 | 0 | 0 |  | 1 | 21 | 1408 | 2206 |
| 1999 | 1 |  | 11 | 240 | 0 | 0 | 0 | 0 |  | 0 | 4 | 11 | 236 |
| 1999 | 2 | 2 | 139 | 513 | 0 | 0 | 7 | 113 |  | 0 | 0 | 132 | 400 |
| 1999 | 3 |  | 508 | 2460 | 284 | 1162 | 188 | 775 |  | 0 | 0 | 36 | 523 |
| 1999 | 4 | 4 | 783 | 1961 | 145 | 757 | 12 | 82 |  | 0 | 0 | 625 | 1123 |
| 1999 |  |  | 1441 | 5174 | 429 | 1919 | 207 | 970 |  | 0 | 4 | 804 | 2282 |
| 2000 | 1 |  | 542 | 872 | 0 | 0 | 0 | 0 |  | 0 | 0 | 542 | 872 |
| 2000 | 2 | , | 277 | 916 | 0 | 0 | 2 | 105 |  | 0 | 0 | 275 | 811 |
| 2000 | 3 |  | 521 | 2277 | 321 | 1448 | 73 | 401 |  | 0 | 24 | 127 | 404 |
| 2000 | 4 | 4 | 196 | 1249 | 192 | 1123 | 4 | 33 |  | 0 | 0 | 0 | 93 |
| 2000 |  |  | 1536 | 5314 | 513 | 2571 | 79 | 539 |  | 0 | 24 | 944 | 2180 |
| 2001 | 1 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |
| 2001 | 2 | 2 | 30 | 497 | 0 | 0 | 29 | 448 |  | 1 | 49 | 0 | 0 |
| 2001 | 3 |  | 720 | 3093 | 661 | 2724 | 58 | 330 |  | 0 | 0 | 1 | 39 |
| 2001 | 4 | 4 | 967 | 2024 | 180 | 934 | 1 | 74 |  | 0 | 0 | 786 | 1016 |
| 2001 |  |  | 1717 | 5614 | 841 | 3658 | 88 | 852 |  | 1 | 49 | 787 | 1055 |
| 2002 | 1 |  | 564 | 423 | 0 | 0 | 0 | 0 |  | 0 | 0 | 564 | 423 |
| 2002 | 2 |  | 734 | 669 | 0 | 0 | 2 | 101 |  | 0 | 0 | 732 | 568 |
| 2002 | 3 |  | 754 | 1885 | 337 | 1371 | 0 | 0 |  | 0 | 0 | 417 | 514 |
| 2002 | 4 | 4 | 134 | 612 | 130 | 610 | 0 | 0 |  | 0 | 0 | 4 | 2 |
| 2002 |  |  | 2186 | 3589 | 467 | 1981 | 2 | 101 |  | 0 | 0 | 1717 | 1507 |

Author: Michael H. Prager; NOAA/NMFS/S.E. Fisheries Science Center
ASPIC User's Manual 101 Pivers Island Road; Beaufort, North Carolina 28516 USA Prager, M. H. 1994. A suite of extensions to a nonequilibrium surplus-production model. Fishery Bulletin 92: 374-389.

CONTROL PARAMETERS USED (FROM INPUT FILE)


NOTE: B1/B $\mathbf{B}_{\text {MY }}$ constraint term contributing to loss. Sensitivity analysis advised.

| Est. B/ $\mathbf{B}_{\text {MSY }}$ coverage index (0 worst, 2 best) : | 0.9419 | < These two measures are defined in Prager |
| :---: | :---: | :---: |
| Est. B/B ${ }_{\text {MSY }}$ nearness index (0 worst, 1 best) : | 0.9419 | $<$ et al. (1996), Trans. A.F.S. 125:729 |

MODEL PARAMETER ESTIMATES (NON-BOOTSTRAPPED)

| Parameter |  | Estimate | Starting guess | Estimated | User guess |
| :---: | :---: | :---: | :---: | :---: | :---: |
| B1R | Starting B/ $\mathbf{B}_{\text {MSY }}$, year 1986 | $2.105 \mathrm{E}+00$ | $1.000 \mathrm{E}+00$ | 1 | 1 |
| MSY | Maximum sustainable yield | $4.869 \mathrm{E}+04$ | $3.000 \mathrm{E}+04$ | 1 | 1 |
| r | Intrinsic rate of increase | $4.008 \mathrm{E}-01$ | $5.000 \mathrm{E}-02$ | 1 | 1 |
| $\begin{aligned} & \cdots(1) \\ & q(1) \end{aligned}$ | Catchability coefficients by fishery: <br> ICE CPUE indices $50 \%$ | $2.227 \mathrm{E}-03$ | $1.000 \mathrm{E}-04$ | 1 | 1 |

MANAGEMENT PARAMETER ESTIMATES (NON-BOOTSTRAPPED)

| Parameter |  | Estimate | Formula | Related quantity |
| :---: | :---: | :---: | :---: | :---: |
| MSY | Maximum sustainable yield | $4.869 \mathrm{E}+04$ | Kr/4 |  |
| K | Maximum stock biomass | $4.859 \mathrm{E}+05$ |  |  |
| $\mathrm{B}_{\text {MSY }}$ | Stock biomass at MSY | $2.430 \mathrm{E}+05$ | K/2 |  |
| $\mathbf{F}_{\text {MSY }}$ | Fishing mortality at MSY | $2.004 \mathrm{E}-01$ | r/2 |  |
| F(0.1) | Management benchmark | $1.804 \mathrm{E}-01$ | $0.9 * \mathbf{F}_{\text {MSY }}$ |  |
| Y (0.1) | Equilibrium yield at F(0.1) | $4.820 \mathrm{E}+04$ | $0.99 *$ MSY |  |
| B. $/ \mathrm{B}_{\text {MSY }}$ | Ratio of $\mathrm{B}(2003)$ to $\mathbf{B}_{\text {MSY }}$ | 1.399E+00 |  |  |
| F. / $\mathbf{F}_{\text {MSY }}$ | Ratio of $\mathrm{F}(2002)$ to $\mathbf{F}_{\text {MSY }}$ | $3.579 \mathrm{E}-01$ |  |  |
| F01-mult | Ratio of $\mathrm{F}(0.1)$ to $\mathrm{F}(2002)$ | $2.515 \mathrm{E}+00$ |  |  |
| Ye./MSY | Proportion of MSY avail in 2003 | $8.405 \mathrm{E}-01$ | $2 * \mathrm{Br}-\mathrm{Br}^{\wedge} 2$ | $\mathrm{Ye}(2003)=4.092 \mathrm{E}+04$ |
| . | Fishing effort at MSY in units | fishery: |  |  |
| $\mathbf{F}_{\text {MSY }}(1)$ | ICE CPUE indices 50\% | $8.998 \mathrm{E}+01$ | r/2q( 1) | $\mathrm{f}(0.1)=8.099 \mathrm{E}+01$ |
| ESTIMATED POPULATION TRAJECTORY (NON-BOOTSTRAPPED) |  |  |  |  |


|  | Year | Estimated total | Estimated starting | Estimated average biomas | Observed total | Model total <br> - | Estimated surplus | Ratio of F mort | Ratio of biomass |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Obs |  |  |  |  |  |  |  |  |  |
| 1 | 1986 | 0.096 | $5.115 \mathrm{E}+05$ | $4.862 \mathrm{E}+05$ | $4.645 \mathrm{E}+04$ | $4.645 \mathrm{E}+04$ | $-2.561 \mathrm{E}+02$ | $4.768 \mathrm{E}-01$ | $2.105 \mathrm{E}+00$ |
| 2 | 1987 | 0.083 | $4.648 \mathrm{E}+05$ | $4.515 \mathrm{E}+05$ | $3.757 \mathrm{E}+04$ | $3.757 \mathrm{E}+04$ | $1.278 \mathrm{E}+04$ | $4.153 \mathrm{E}-01$ | $1.913 \mathrm{E}+00$ |
| 3 | 1988 | 0.073 | $4.400 \mathrm{E}+05$ | $4.333 \mathrm{E}+05$ | $3.143 \mathrm{E}+04$ | $3.143 \mathrm{E}+04$ | $1.881 \mathrm{E}+04$ | $3.620 \mathrm{E}-01$ | $1.811 \mathrm{E}+00$ |
| 4 | 1989 | 0.131 | $4.274 \mathrm{E}+05$ | $4.120 \mathrm{E}+05$ | $5.391 \mathrm{E}+04$ | $5.391 \mathrm{E}+04$ | $2.508 \mathrm{E}+04$ | $6.530 \mathrm{E}-01$ | $1.759 \mathrm{E}+00$ |
| 5 | 1990 | 0.113 | $3.985 \mathrm{E}+05$ | $3.913 \mathrm{E}+05$ | $4.420 \mathrm{E}+04$ | $4.420 \mathrm{E}+04$ | $3.053 \mathrm{E}+04$ | $5.638 \mathrm{E}-01$ | $1.640 \mathrm{E}+00$ |
| 6 | 1991 | 0.185 | $3.848 \mathrm{E}+05$ | $3.678 \mathrm{E}+05$ | $6.788 \mathrm{E}+04$ | $6.788 \mathrm{E}+04$ | $3.577 \mathrm{E}+04$ | $9.211 \mathrm{E}-01$ | $1.584 \mathrm{E}+00$ |
| 7 | 1992 | 0.185 | $3.527 \mathrm{E}+05$ | $3.409 \mathrm{E}+05$ | $6.310 \mathrm{E}+04$ | $6.310 \mathrm{E}+04$ | $4.074 \mathrm{E}+04$ | $9.237 \mathrm{E}-01$ | $1.452 \mathrm{E}+00$ |
| 8 | 1993 | 0.236 | $3.304 \mathrm{E}+05$ | $3.146 \mathrm{E}+05$ | $7.420 \mathrm{E}+04$ | $7.420 \mathrm{E}+04$ | $4.439 \mathrm{E}+04$ | $1.177 \mathrm{E}+00$ | $1.360 \mathrm{E}+00$ |
| 9 | 1994 | 0.297 | $3.006 \mathrm{E}+05$ | $2.814 \mathrm{E}+05$ | $8.357 \mathrm{E}+04$ | $8.357 \mathrm{E}+04$ | $4.738 \mathrm{E}+04$ | $1.482 \mathrm{E}+00$ | $1.237 \mathrm{E}+00$ |
| 10 | 1995 | 0.214 | $2.644 \mathrm{E}+05$ | $2.606 \mathrm{E}+05$ | $5.574 \mathrm{E}+04$ | $5.574 \mathrm{E}+04$ | $4.843 \mathrm{E}+04$ | $1.067 \mathrm{E}+00$ | $1.088 \mathrm{E}+00$ |
| 11 | 1996 | 0.161 | $2.571 \mathrm{E}+05$ | $2.605 \mathrm{E}+05$ | $4.186 \mathrm{E}+04$ | $4.186 \mathrm{E}+04$ | $4.843 \mathrm{E}+04$ | $8.019 \mathrm{E}-01$ | $1.058 \mathrm{E}+00$ |
| 12 | 1997 | 0.162 | $2.637 \mathrm{E}+05$ | $2.663 \mathrm{E}+05$ | $4.305 \mathrm{E}+04$ | $4.305 \mathrm{E}+04$ | $4.823 \mathrm{E}+04$ | $8.066 \mathrm{E}-01$ | $1.085 \mathrm{E}+00$ |
| 13 | 1998 | 0.142 | $2.688 \mathrm{E}+05$ | $2.735 \mathrm{E}+05$ | $3.889 \mathrm{E}+04$ | $3.889 \mathrm{E}+04$ | $4.791 \mathrm{E}+04$ | $7.096 \mathrm{E}-01$ | $1.107 \mathrm{E}+00$ |
| 14 | 1999 | 0.123 | $2.779 \mathrm{E}+05$ | $2.842 \mathrm{E}+05$ | $3.499 \mathrm{E}+04$ | $3.499 \mathrm{E}+04$ | $4.727 \mathrm{E}+04$ | $6.144 \mathrm{E}-01$ | $1.144 \mathrm{E}+00$ |
| 15 | 2000 | 0.127 | $2.901 \mathrm{E}+05$ | $2.948 \mathrm{E}+05$ | $3.747 \mathrm{E}+04$ | $3.747 \mathrm{E}+04$ | $4.647 \mathrm{E}+04$ | $6.342 \mathrm{E}-01$ | $1.194 \mathrm{E}+00$ |
| 16 | 2001 | 0.072 | $2.991 \mathrm{E}+05$ | $3.107 \mathrm{E}+05$ | $2.252 \mathrm{E}+04$ | $2.252 \mathrm{E}+04$ | $4.487 \mathrm{E}+04$ | $3.618 \mathrm{E}-01$ | $1.231 \mathrm{E}+00$ |
| 17 | 2002 | 0.072 | $3.215 \mathrm{E}+05$ | $3.311 \mathrm{E}+05$ | $2.374 \mathrm{E}+04$ | $2.374 \mathrm{E}+04$ | $4.226 \mathrm{E}+04$ | 3.579E-01 | $1.323 \mathrm{E}+00$ |
| 18 | 2003 |  | $3.400 \mathrm{E}+05$ |  |  |  |  |  | $1.399 \mathrm{E}+00$ |

Table 9.2.2

RESULTS FOR DATA SERIES \# 1 (NON-BOOTSTRAPPED) ICE CPUE indices $50 \%$

| Data type CC: CPUE-catch series |  |  |  |  |  |  | Series weight: 1.000 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Observed | Estimated | Estim | Observed | Model | Resid in | Resid in |
| Obs | Year | CPUE | CPUE | F | yield | yield | $\log$ scale | $\log$ yield |
| 1 | 1986 | $1.000 \mathrm{E}+03$ | $1.083 \mathrm{E}+03$ | 0.0955 | $4.645 \mathrm{E}+04$ | $4.645 \mathrm{E}+04$ | 0.07949 |  |
| 2 | 1987 | $1.090 \mathrm{E}+03$ | $1.005 \mathrm{E}+03$ | 0.0832 | $3.757 \mathrm{E}+04$ | $3.757 \mathrm{E}+04$ | -0.08074 |  |
| 3 | 1988 | $9.890 \mathrm{E}+02$ | $9.649 \mathrm{E}+02$ | 0.0726 | $3.143 \mathrm{E}+04$ | $3.143 \mathrm{E}+04$ | -0.02472 |  |
| 4 | 1989 | $9.460 \mathrm{E}+02$ | $9.174 \mathrm{E}+02$ | 0.1309 | $5.391 \mathrm{E}+04$ | $5.391 \mathrm{E}+04$ | -0.03068 |  |
| 5 | 1990 | $9.040 \mathrm{E}+02$ | $8.714 \mathrm{E}+02$ | 0.1130 | $4.420 \mathrm{E}+04$ | $4.420 \mathrm{E}+04$ | -0.03676 |  |
| 6 | 1991 | $9.070 \mathrm{E}+02$ | $8.190 \mathrm{E}+02$ | 0.1846 | $6.788 \mathrm{E}+04$ | $6.788 \mathrm{E}+04$ | -0.10208 |  |
| 7 | 1992 | $7.050 \mathrm{E}+02$ | $7.592 \mathrm{E}+02$ | 0.1851 | $6.310 \mathrm{E}+04$ | $6.310 \mathrm{E}+04$ | 0.07408 |  |
| 8 | 1993 | $6.360 \mathrm{E}+02$ | $7.006 \mathrm{E}+02$ | 0.2358 | $7.420 \mathrm{E}+04$ | $7.420 \mathrm{E}+04$ | 0.09674 |  |
| 9 | 1994 | $5.690 \mathrm{E}+02$ | $6.267 \mathrm{E}+02$ | 0.2970 | $8.357 \mathrm{E}+04$ | $8.357 \mathrm{E}+04$ | 0.09653 |  |
| 10 | 1995 | $5.830 \mathrm{E}+02$ | $5.803 \mathrm{E}+02$ | 0.2139 | $5.574 \mathrm{E}+04$ | $5.574 \mathrm{E}+04$ | -0.00461 |  |
| 11 | 1996 | $5.880 \mathrm{E}+02$ | $5.801 \mathrm{E}+02$ | 0.1607 | $4.186 \mathrm{E}+04$ | $4.186 \mathrm{E}+04$ | -0.01359 |  |
| 12 | 1997 | $6.170 \mathrm{E}+02$ | $5.931 \mathrm{E}+02$ | 0.1616 | $4.305 \mathrm{E}+04$ | $4.305 \mathrm{E}+04$ | -0.03947 |  |
| 13 | 1998 | $6.210 \mathrm{E}+02$ | $6.091 \mathrm{E}+02$ | 0.1422 | $3.889 \mathrm{E}+04$ | $3.889 \mathrm{E}+04$ | -0.01940 |  |
| 14 | 1999 | $6.610 \mathrm{E}+02$ | $6.329 \mathrm{E}+02$ | 0.1231 | $3.499 \mathrm{E}+04$ | $3.499 \mathrm{E}+04$ | -0.04344 |  |
| 15 | 2000 | $6.960 \mathrm{E}+02$ | $6.565 \mathrm{E}+02$ | 0.1271 | $3.747 \mathrm{E}+04$ | $3.747 \mathrm{E}+04$ | -0.05840 |  |
| 16 | 2001 | $6.780 \mathrm{E}+02$ | $6.918 \mathrm{E}+02$ | 0.0725 | $2.252 \mathrm{E}+04$ | $2.252 \mathrm{E}+04$ | 0.02019 |  |
| 17 | 2002 | $6.760 \mathrm{E}+02$ | $7.373 \mathrm{E}+02$ | 0.0717 | $2.374 \mathrm{E}+04$ | $2.374 \mathrm{E}+04$ | 0.08682 |  |



SMEN cpue 86 to 02
Observed (O) and Estimated (*) CPUE for Data Series \# 1 -- ICE CPUE indices 50\%


Table 9.2.2 continued.


Table 9.2.3 Deep-sea $S$. mentella. Retrospective analysis from a ASPIC stock production model. The table shows MSY, K, r , and $\mathrm{K} / \mathrm{r}$ for different length of the input series using CPUE data from both the Faroese and Icelandic trawl fleet.

| Period | MSY | K | r | K/r |
| :--- | ---: | ---: | ---: | ---: |
| $1986-2002$ | 48,690 | 485,900 | 0.40 | 1212325 |
| $1986-2001$ | 50,110 | 426,600 | 0.47 | 908046 |
| $1986-2000$ | 51,460 | 382,200 | 0.54 | 709485.8 |
| $1986-1999$ | 51,850 | 371,500 | 0.56 | 665412.9 |
| $1986-1998$ | 52,750 | 350,600 | 0.60 | 582488.8 |
| $1986-1997$ | 55,430 | 296,800 | 0.75 | 397269.4 |
| $1986-1996$ | 58,150 | 249600 | 0.93 | 267811.2 |

Table 9.2.4 Deep-sea $S$. mentella. Catch projection for 2004. The table gives the nominal catch (tonnes), CPUE, and fishing effort Va and Vb , the catch for the 2004 fishing year in Va and Vb based on mean CPUE and effort of 2000-2002, and total catch in these two areas.

|  | Iceland |  |  | Faroese Islands |  |  |
| :--- | ---: | :---: | :---: | ---: | :---: | ---: |
| Year | Landings | CPUE | Effort | Landings | CPUE | Effort |
| 1986 | 18,898 | 1000 | 19 |  |  |  |
| 1987 | 19,293 | 1090 | 18 |  |  |  |
| 1988 | 14,290 | 989 | 14 |  |  |  |
| 1989 | 40,269 | 946 | 43 |  |  |  |
| 1990 | 28,429 | 904 | 31 |  | 501 | 26 |
| 1991 | 47,651 | 907 | 53 | 12,897 | 384 | 33 |
| 1992 | 43,414 | 705 | 62 | 12,533 | 321 | 24 |
| 1993 | 51,221 | 636 | 81 | 7,801 | 308 | 22 |
| 1994 | 56,720 | 569 | 100 | 6,899 | 311 | 18 |
| 1995 | 48,708 | 583 | 84 | 5,670 | 292 | 18 |
| 1996 | 34,741 | 588 | 59 | 5,337 | 292 | 14 |
| 1997 | 37,876 | 617 | 61 | 4,558 | 337 | 14 |
| 1998 | 33,125 | 621 | 53 | 4,089 | 295 | 17 |
| 1999 | 28,590 | 661 | 43 | 5,294 | 307 | 17 |
| 2000 | 30,696 | 696 | 44 | 4,841 | 289 | 17 |
| 2001 | 17,313 | 678 | 26 | 4,247 | 308 | 14 |
| 2002 | 19,148 | 676 | 28 | 2,674 | 373 | 7 |
| Av. 00-02 |  | $\mathbf{6 8 3}$ | $\mathbf{3 3}$ |  | $\mathbf{3 2 3}$ | $\mathbf{1 3}$ |


| Catch 2004 |  |  |  |
| :--- | :---: | :---: | :---: |
| $=$Mean CPUE(2000-2002)* <br> Effort(2000-2002) | Mean |  |  |
| Level | Va | Vb | Total |
| $120 \%$ | 26,777 | 4,878 | 31,656 |
| $100 \%$ | 22,314 | 4,065 | 26,380 |
| $75 \%$ | 16,736 | 3,049 | 19,785 |
| $50 \%$ | 11,157 | 2,033 | 13,190 |



Figure 9.1.1 Deep-sea $S$. mentella. Catch in Icelandic waters 2000-2002.


Figure 9.1.2
Redfish (both S. mentella and S. marinus) catch in Icelandic waters 2000-2002 (all years combined) as recorded by log-books. The map also shows the line used by the Icelandic authorities to separate the landing statistics between deep-sea $S$. mentella and pelagic deep-sea $S$. mentella. The catches west of the "redfish line" is from the pelagic fishery, whereas the catch north and east of the line is of "shelf type".


Figure 9.1.3 Deep-sea S. mentella. Nominal catch (tonnes) by month in Icelandic waters 1998-2003.


Figure 9.1.4 Deep-sea $S$. mentella. Length distributions from the Icelandic bottom trawl catch and landings in Division Va 1989-2002.


Figure 9.1.5 Deep-sea $S$. mentella. Length distribution from the Faroese fleet catch and landings in Division Vb 2001 and 2002.


Figure 9.1.6 Deep sea S. mentella. Length distribution of the German fleet catches in Division XIV 1999-2002.


Figure 9.2.1 Deep-sea $S$. mentella. CPUE, relative to 1986, from the Icelandic bottom trawl fishery in Division Va. CPUE based on a GLM model, based on data from log-books and where at least $50 \%$ of the total catch in each tow was deep-sea $S$. mentella. Also shown is fishing effort (hours fished in thousands).


Figure 9.2.2 Deep-sea $S$. mentella.. CPUE (kg/hour) and fishing effort (in thousands) from the Faroese CUBA fleet 1991-2002 and where 70\% of the total catch was deep-sea S. mentella .


Figure 9.2.3 Deep-sea $S$. mentella. CPUE (kg/hour) from the Faroese summer survey 1996-2002.


Figure 9.2.4 Deep-sea $S$. mentella $(15-35 \mathrm{~cm})$ on the continental shelf. Length composition off Greenland and Iceland as derived from the German and Icelandic groundfish surveys, 1985-1994.





Figure 9.2.4 Continued.


Figure 9.2.5 Deep-sea $S$. mentella $(>=17 \mathrm{~cm})$ on the continental shelf. Survey abundance indices for East and West Greenland and Iceland derived from the German and Icelandic groundfish surveys, 19852002.


Figure 9.2.6 Deep-sea $S$. mentella ( $>=17 \mathrm{~cm}$ ) on the continental shelf. Survey biomass indices for East and West Greenland and Iceland derived from the German and Icelandic groundfish surveys, 19852002.



Figure 9.2.7 Deep-sea S. mentella. Retrospective analysis of the ASPIC model for F./F $\mathbf{F}_{\text {MSY }}$ (a) and B./B $\mathbf{B}_{\mathrm{MSY}}$ (b).

This section includes information on the pelagic fishery for $S$. mentella both shallower and deeper than 500 m in the Irminger Sea (Subarea XII, parts of Division Va, Subarea XIV and eastern parts of NAFO Divisions 1F, 2H and 2J).

Under chapter 7.5, comments are made on special requests in the ToR. Aside from what is said there, the WG refers to last years' reports on the matter of stock identification in the area.

### 10.1 Fishery

### 10.1.1 Summary of the development of the fishery

Russian trawlers started fishing pelagic S. mentella in 1982. Vessels from Bulgaria, the former GDR and Poland joined those from in 1984. Total catches increased from 60600 t in 1982 to 105000 t . in 1986. Since 1987, the total landings decreased to a minimum in 1991 of 28000 t mainly due to effort reduction. Since 1989, the number of countries, participating in the pelagic $S$. mentella fishery gradually increased. As a consequence, total catches also increased after the 1991 minimum and reached a historical high of 180000 t in 1996 (Tables 10.1.1-10.1.2). Since 2000, the WG estimate of the catch has been between 126000 and 132000 t . This is probably an underestimate due to poor reporting of catches from non-ICES countries.

In the period 1982-1992, the fishery was carried out mainly from April to August. In 1993-1994, the fishing season was prolonged considerably, and in 1995 the fishery was conducted from March to December. Since 1997, the main fishing season occurred during the second quarter. The pattern in the fishery has been reasonably consistent in the last 5 years and can be described as follows: In the first months of the fishing season (which usually starts in early April), the fishery is conducted in area west of $32^{\circ} \mathrm{W}$ and north of $61^{\circ} \mathrm{N}$; in May and June the fishery is more or less at same areas, but in July (August), the fleet moves to areas south of $60^{\circ} \mathrm{N}$ and west of about $32^{\circ} \mathrm{W}$ where the fishery continues until October (see figures 7.5.1-7.5.6). There is very little fishing activity in the period from November until late March or early April when the next fishing season starts. The fleets participating in this fishery have continued to develop their fishing technology, and most trawlers now use large pelagic trawls ("Gloria"-type) with vertical openings of $80-150 \mathrm{~m}$. The vessels have operated in 1998-2002 at a depth range of 200 to 950 m , but mainly deeper than 600 m in the first and second quarter but at depths shallower than 500 m in third and fourth quarter. Discard is not considered to be significant for this fishery (see 10.1.3).

The following text table summarises the available information from fishing fleets in the Irminger Sea in 2002:

| Faroes | 3 factory trawlers |
| :--- | :---: |
| Germany | 7 factory trawlers |
| Greenland | 1 factory trawler |
| Iceland | 25 factory trawlers and 2 freshfish trawlers |
| Norway | 4 factory trawlers |
| Russia | 29 factory trawlers |
| Portugal | 6 factory trawlers |
| Spain | 6 factory trawlers |

A summary of the catches by depth by nation as estimated by the Working Group is given in Table 10.1.2.

### 10.1.2 Description on the fishery of various fleets

### 10.1.2.1 Faroes

The Faroese fishery for pelagic redfish in the Irminger Sea and adjacent waters started in 1986. In the first years, only 12 trawlers participated in the fishery. Fishing depths were mainly shallower than 500 m although some trials were made down to about 700 m . From 1994 onwards, several trawlers have made trips to this area fishing almost exclusively deeper than 500-600 m .

Since 1999 the Faroese fishery started in international waters in the NE part of the Irminger Sea in the middle/late April (ICES Division XIV). Up to late July, the fishing area was mainly between $61^{\circ} \mathrm{N}-62 \mathrm{~N}$ and $27^{\circ} \mathrm{N}-30^{\circ} 00^{\prime} \mathrm{W}$, then they moved to the SW, to south of $60^{\circ} \mathrm{N}$ and west of $38^{\circ} \mathrm{W}$ (ICES Division XII), fishing mostly within the Greenlandic EEZ.

Three trawlers participated in 2002. The fishing depth from the beginning of the fishery to July was nearly exclusively deeper than 600 m , but from July onwards, the fish was taken at shallower depths than 600 m .

### 10.1.2.2 Germany

The reported effort in 2002 is the lowest observed in the last eight years and amounted to 12700 hours only. As observed in previous years, the majority of the 2002 effort was applied during the second and third quarters. During the second quarter in 2002, the hauls were almost exclusively distributed in NEAFC Regulatory Area of ICES Division XIV between the Greenland and Icelandic EEZs. In 2002, there was significant fishing effort exerted in the NAFO Subarea 1F mainly within the NAFO Regulatory Area. The decrease of annual landings discontinued in 2002 with a catch figure of 10700 tons in 2001. In 2002, $18 \%$ or 2300 tons of the total landings were taken in the NAFO Div. 1 F . During 1995-1999, the overall unstandardised CPUE decreased from $2055 \mathrm{~kg} / \mathrm{h}$ by $53 \%$ to $970 \mathrm{~kg} / \mathrm{h}$. In 2000-2002, the CPUE remained at that low level. Given the technical, temporal, geographical and depth changes of the fishing activities the relevance of the estimated reduction in CPUE as indicator of stock abundance remained difficult to assess.

### 10.1.2.3 Greenland

The Greenlandic fleet was fishing in the same area as the Icelandic fleet (see below), and therefore, the greenlandic logbook data were included in the figure of the Icelandic fishery.

### 10.1.2.4 Iceland

Catches in 1995-2000 were generally taken in the area between the Greenlandic EEZ and the Reykjanes Ridge. Since 1996, the catches have mostly been taken close to or inside the 200-mile boundary Southwest of Iceland. In recent years, the fishery has started in April close to the Icelandic 200-mile boundary and then moved northward in May-July. In the springtime and until June, the largest proportions of the catches were taken deeper than 500 m . In 1998, the fishery expanded further north in July-September. In 1999, a similar pattern was observed, except that the fishery did not continue close to the shelf of Iceland. The few vessels that had quota left after that, moved about 480 nautical miles to Southwest, to the area S-SE of Cape Farewell (Subarea XII), where they fished shallower than 500 m depth in July-September. In 2000, the fishery started in April at the same locations as in the past and moved slowly northward until the fishery ended in July due to quota limitation. The Icelandic trawlers fished mainly at a depth of 600-800 m during the period 1995-2000 (Figure 10.1.1). In 2000 , less than $8 \%$ of the catches in the log-books were reported shallower than 500 m depth and no catches were reported at depths shallower than 400 m . In 2001-2002, the fishery started in late April and until middle of July, the fishery was nearly exclusively within the Icelandic EEZ moving slowly in northward direction. In May - July over $90 \%$ of the catches were taken at depths deeper than 600 m . From the middle of July until the end of the fishing season the fishery continued in the area Southeast of Cape Farwell, mostly between 38 and $42^{\circ} \mathrm{W}$. Only about $12 \%$ of the Icelandic catches in 2002 were taken in the "south-western" fishing area shallower than 400 m depth. Length distribution from the catch is shown in Figure 10.1.2.

### 10.1.2.5 Norway

Information on the fishery in 1998 and 1999 indicated a depth shift in the fishery, from fishing $95 \%$ of its catch shallower than 500 m in 1998 to fishing exclusively deeper than 500 m in 1999. The catches in 1999 were taken in areas XII and XIV from April to August, at a ratio of about 2:3. In 2000, Norway fished 6075 t whereof 3823 t were taken in ICES Subarea XIV and 2252 t in Subarea XII. The fishing season was from April - September. In 2001 and 2002, the fishery started in April, close to the Icelandic 200 miles boundary (Subarea XIV). The fishery continued there until beginning of June and nearly $80-85 \%$ of the total catch was caught below 600 m . Then the fleet moved to Subarea XII between 55 and $58^{\circ} \mathrm{N}$ and between 40 and $42^{\circ} \mathrm{W}$. There is no information on the trawling depth or about length distributions in the catches.

### 10.1.2.6 Russia

The regular Russian commercial fishery for pelagic redfish in the Irminger Sea started in 1982. Total catch of redfish taken by the USSR/Russia makes up about 0,76 mill. $t$ or $43 \%$ of the total world catch for a whole period of the fishery in the Irminger Sea. In 1982-1988, the annual Russian catch of redfish constituted 60-85 thou. t. The fishery duration was 4-4,5 months. In 1989-1994, the catch decreased to 9-25 thou. t. Fishing efficiency of STM-type vessels was 10-15 $t$ per a vessel/fishing day. A shift of the fishery to the depths deeper than 500 m , and due to an increase in trawl size, an increase in fishing efficiency was observed in 1994. A reduction in redfish catches from the depths deeper than 500 m has been observed since 1997. The extension of fishing period to 8 months and extension of areas due to the increased fishery within the 200-mile zone of Greenland and adjacent areas of the Labrador Sea occurred simultaneously.

In 2002, Russian fishery for pelagic Sebastes mentella in the Irminger Sea and NAFO Div. 1F lasted from April to October. It was conducted by 29 trawlers of different types. In April, the fishery for spawning concentrations of redfish started on traditional fishing grounds near the Icelandic Economic Zone at $600-900 \mathrm{~m}$ depth. During the second quarter of the year the redfish fishery took place in the open part of ICES Subarea XIV where CPUE was $825 \mathrm{~kg} / \mathrm{h} .51 \%$ of total catch and $52 \%$ of fishing effort were recorded in June-July. In the third quarter, following the feeding migration of redfish, the fishery shifted south-westwards to ICES Subarea XII and the fishery zone of Greenland (200-650 m in depth). At the same time one Russian vessel conducted successful fishery in Subarea XIV ( $62^{\circ} 00^{\prime} \mathrm{N} 28^{\circ} 00^{\prime} \mathrm{N}-30^{\circ} 00^{\prime}$ W) at 600-850 m depth until middle of September. In late July Russian vessels started redfish fishery in NAFO Div.1F at 250-400 m depth. In July-September CPUE in Div. 1F amounted to $1185 \mathrm{~kg} / \mathrm{h}$. The total Russian catch of redfish in 2002 is estimated to be 27349 t in ICES Subareas XII and XIV and 4820 t in NAFO Div.1F (WD 7).

### 10.1.2.7 Spain

Four Spanish freezer trawlers fished pelagic redfish in 1995-1997, increasing to 6 vessels since 1998. The fleet has used a Gloria-type pelagic trawls, with a maximum vertical opening of $80-120 \mathrm{~m}$. The fishery in ICES Subareas XII and XIV shows a persistent seasonal pattern in terms of its geographical and depth distribution. The main fishing occurs in the second and third quarter each year. In the second quarter, the fishery takes place in Subarea XIV between the Greenland and Iceland EEZs deeper than 500 m , capturing fish of bigger size. The proportion of females in the catches is greater than for the males. The yields obtained in this quarter are larger and the mean trawling time of the hauls is shorter than in the third quarter. In the third quarter, the fleet moves towards the Southwest to ICES Subarea XII, and the depth of hauls is shallower than 500 meters. The length distributions in the catches are lower than in the second quarter and are unimodal and relatively stable in time. The proportion of males in the catches is higher than for the females. The yields are smaller and the mean trawling time of the hauls is greater than those of the second quarter.

### 10.1.2.8 Portugal

The Portuguese fleet commenced the fishery in 1994. In 2002, six trawlers participated, fishing with a large pelagic net. Based on the observed vessel and observed period (Div. XII in September and October from 203m to 380 m depths, and in Div. XIVb from May to July from 400 m to 770 m depths), the months with highest catch rates were October in Div. XII, and June in Div. XIVb. Despite the concentration of fishing effort on Div. XIVb, higher CPUE was recorded in Div.XII than in the northern Div. XIV b. In Div. XIVb CPUE at depth shallower than 500 m were smaller than those at greater depth were.

In Div. XII, lengths between 32 cm and 36 cm dominated catches (mean length and weight of 34 cm and 555 g ). In Div. XIVb at depth $<500 \mathrm{~m}$, lengths between 39 cm and 42 cm dominated catches (mean length and weight of 39 cm and 815 g ). At depth $>500 \mathrm{~m}$ and for the all Div.XIVb, lengths between 38 cm and 43 cm dominated catches (mean length and weight of 39 cm and 838 g ).

### 10.1.2.9 Other nations

No information on the fishing areas, seasons and depths of the fleets of other nations was available for the Working Group.

### 10.1.3 Discards

Icelandic landings of oceanic redfish were raised by $16 \%$ prior to 1996 to into account discards of redfish infested with Sphyrion lumpi. This value of was based on measurements from 1991-1993 when the fishery was mostly on depths shallower than 500 m . In May-July 1997, discard measurements on 10 vessels showed a discard rate of $10 \%$. This was added to the landings in 1996 and 1997. A new measurement from 1998 shows that the discard rate has decreased to $2 \%$. Information from observers from 2000-2002 indicate that discards is negligible, and therefore no catches were added to the Icelandic landings.

Norwegian fishermen currently report approximately $3 \%$ discards of redfish infested with parasites. This percentage has in recent years become less due to a change in the production from Japanese cut to mainly fillets at present.

The Spanish discards are based on measurements made by the scientific observers. Discard of the Spanish fleet were often composed of fish infested with Sphyrion lumpi. In 1995, about $4 \%$ of the total catches were discarded, while in 1996, it was $6.5 \%$. From 1997-2000, the discards of the Spanish fleet have been insignificant. In 2001, $4.4 \%$ of the total catches were discarded. This variability in the discard is also observed at different depths. The discarded percentage being much larger at depth greater than 500 m . Since 1997, this variability can be due to that the percentage of discards does not depend directly on parasite fish by Sphyrion lumpi, but it is related with the haul catch. When the
haul catch is very high, the fish is discarded under worse conditions by the lack of time to elaborate the whole catch. When the catches are between the standard values there is enough time to elaborate the whole fish, even the infested ones, and there are no discards.

The level of redfish discarded by the Portuguese fleet, based on the observer reports, has been very small, between 0.6 and $1.0 \%$ of the catch.

No information on possible discards was available from other countries participating in this fishery.

### 10.1.4 Trends in landings

Total catches in 2002 is estimated to be about 132000 t , similar as in 2001 . The catch estimates for 2002 might increase due to the lack of reporting from some countries participating in the fishery. Catches from the beginning of the fishery is given it table 10.1.1-10.1.2

At the beginning of the fishery in 1982, catches of pelagic redfish were reported from both Subareas XII and XIV. Most of the catches were taken in Subarea XII (40 000-60 000 t) prior to 1985, and then the greater part of the catches was reported from Subarea XIV. The landings from Subarea XII were again in the majority in 1994 and in 1995 with 94000 t and 129000 t landed respectively. In 1996-1999, the main part of the total catch was taken from Div. Va and Subarea XIV (Table 10.1.1). In 2000, considerable amounts of the catches were taken in NAFO Div. 1F, as observed in this magnitude for the first time. In 2001 and 2002 about 6791 and 7639 t of pelagic $S$. mentella were reported in NAFO Div 1F, 2H and 2J, respectively.

Pelagic S. mentella fishery in ICES Div. Va started in 1992. The catch varied from 2 000-14 000 from 1992-1995. From 1995-2000, the catches in Div. Va increased to about 45000 t (Table 10.1.1). Total catches in 2001 and 2002 were 28000 and 37000 t respectively.

### 10.1.5 Biological sampling from the fishery

Length distributions of pelagic S. mentella from German, Icelandic, Russian, Portugese and Spanish commercial catches were reported for 2002 and are given in Figure 10.1.2. A bimodal distribution over the past years could be observed as a reflection of the samples being taken from different areas. Figure 10.1.3 illustrates the depth effect on the length distributions in the German and Spanish catches, taken shallower and deeper than 500 m .

The 2002 biological sampling from catches and landings of pelagic $S$. mentella in each Subarea/Division and by gear type is shown in the text table below.

| Country | Area | Gear | Landings $(\mathrm{t})$ | No. of samples | No. of fish measured |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Germany | XII, XIV and NAFO 1F | Pelagic | 13191 | 33 | $?$ |
| Iceland | XIV and Va | Pelagic | 44430 | 116 | 16371 |
| Russia | XII, XIV and NAFO 1F | Pelagic | 32169 | 473 | 152202 |
| Spain | XII, XIV | Pelagic | 8950 | 57 | 11262 |
| Portugal | XII, XIV | Pelagic | 3164 | 87 | 6960 |

The catches in 1999 and 2000, and also the acoustic survey in 1999, suggested that a new cohort is entering into the fishable stock of pelagic redfish.

Age readings within an otolith exchange between Germany, Iceland and Norway, based on material collected in July 1999 (ICES 2002 NWWG -WD9), showed that this cohort is mainly consisting of 10 year old fish and that ageing error for fish older than 20 years is relatively high. If agreement is defined as $\pm 5$ years, approximately $90 \%$ agreement would be obtained. A second set of age reading results within an otolith exchange program between Germany, Iceland, Norway and Spain based on material collected in 1998 and 1999 (WD11), shows the same results.

### 10.2.1 Survey data

There were no surveys conducted in 2002. The main results of the 2001 trawl-acoustic survey (ICES CM 2002/D:08 Ref.ACFM). are described in the report of the NWWG in 2002 and the results are given in Table 10.2.1-10.2.2. There will be new survey in June/July 2003 with participation of Russia, Germany and Iceland (ICES CM 2003/D:2).

### 10.2.2 CPUE

Non standardised CPUE (Table 10.2.3 and Figure 10.2.3), series for Bulgarian, German, Icelandic, Spanish, Norwegian and Russian fleets are given. Figures 10.2.3.a and 10.2.3.b show the overall CPUE from different fleets in recent years, in depths shallower and deeper than 500 m , respectively. In Figure 10.2.3.a, along with estimated biomass derived from the international and Russian hydroacoustic surveys. In recent years, there is no trend in CPUE, both shallower and deeper than 500 m (Figures 10.2.3.a-10.2.3.b).

Standardised CPUE (Figure 10.2.3.c), derived from a GLM CPUE model incorporating data from Germany (1995-2002), Iceland (1995-2002), Greenland (1999-2002), Faroe Island (1995-2001), Russia (1997-2001) and Norway (1995-2002) is given. The model takes into account year, month, vessel and area (ICES statistical square). The model was run on as desegregated data as possible from a joint database (WD 18) and the outcomes of 3 model runs are given in Table 10.2.4. The model shows that the index is fluctuating both for the south-western and northeastern fishing area. The value in 2002 has increased for the northeastern part but remains similar for the southwestern area, compares with previous yea. Overall, the GLM model indicates a relatively stable CPUE since 1995 both shallower and deeper than 500 m . The minor changes seen in the series, compared with the run from last year, are because data from new nations have been added to the database (Russia 1997-2001 and Faroe Island form 1995-2001).

### 10.2.3 Ichthyoplankton assessment

The traditional ichthyoplankton survey, conducted by Russia in 1982-1995 has not been carried out since 1996. The historical series of ichthyoplankton surveys was presented in the 2000 Working Group report.

### 10.2.4 State of the stock

Table 10.2.1 shows available survey estimates of stock size by acoustic and trawls. The biomass can be estimated acoustically for depths less than 500 m . Acoustic biomass estimates were relatively stable during 1991 to 1995, but they have declined substantially, from 2.48 million tonnes in 1995 to 0.72 million tonnes in 2001. The acoustic estimates from the last three surveys are considered minimum biomass estimates because trawl sets during those surveys have shown that there was considerable redfish biomass deeper than the depths where biomass can be estimated acoustically. However the proportion of fish above and below 500 m is not known to be stable over years and it cannot be concluded that acoustic biomass estimates prior to 1996 are minimum biomass estimates, because of high variances in the acoustic surveys for those years. These possible changes in the depth distribution above and below 500 m combined with the differences in geographic coverage in different years mean that the acoustic biomass series cannot be interpreted as a consistent series showing relative changes in stock size. It is not known if the trawl survey biomass estimates are minimum or if they can overestimate stock size.

Adding the trawl biomass estimate below 500 m to t he acoustic estimates ( 1.8 million tonnes) or adding the two trawl biomass estimates together ( 2.1 million tonnes) indicates that the biomass in 2001 is probably in the order of 2 million tonnes, distributed also in large portions of the NAFO Convention Area down to depth of 1000 m .

Available CPUE series show that the pelagic redfish CPUE has remained stable since 1995 for all fishing areas as well as separated above and below 500 m depth. There are great seasonal, geographical and depth changes of the fishing activities and the fishery is on schooling aggregations. Therefore CPUE series might not indicate or reflect actual status of the stocks and might thus be to optimistic. Comparing figures of the fishery in recent years (Figure 7.5.1-7.5.4) with the distribution from the surveys (Figure 10.2.2) it can be seen that the fish accumulates in fishable concentrations in relatively small area, compared with the distribution area.

Taking into account the uncertainty in stock indicators, it is not known if the exploitation rate generated by recent catches is above or below $5 \%$ exploitation rate which has been suggested suitable for such a long lived species.

Based on all the available data, the recent exploitation level seems not to cause stock size reduction.

The former proposed MBAL biomass reference of 1.5 Mill. t is considered inappropriate as it was derived from a production model disregarding the increased knowledge about the stock distribution and expanded fishing grounds. None of the available data series are considered appropriate to develop reference.

### 10.4 Management considerations

The working group had again difficulties in obtaining catch estimates from the various international fleets like in the past.

An update on the pelagic fishery, in particular with respect to seasonal and area distribution, was requested in the ToR. Catch rates shallower than 500 m remained steady but low, and deeper than 500 m remained steady. The main new feature of the fishery in recent years is a clear distinction between two widely separated grounds fished at different seasons and different depths. Since 2000 the more southwesterly fishing ground extended also into the NAFO Convention Area. The parameters analysed so far do suggest, however, that the newly discovered aggregations in the NAFO Convention Area do not form a separate stock component. NAFO Scientific Council does agree with this conclusion. No new survey results were available to the working group, but the 1999 and 2001 surveys indicated that about one third of the stock is distributed in the NAFO Convention Area.. The genetic structure of the pelagic and demersal stocks of deep-sea redfish (S. mentella) in the North Atlantic remains poorly known, but further research is currently being carried out.

There will be a survey on pelagic redfish in 2003 and the report will be available in September 2003.
Considering the uncertainty related to definition of stock units, action must be taken in accordance with the precautionary approach and attempts be made to manage each stock component separately until better knowledge on the relationship among units are known. Given the current fishing pattern (the deep water fishery in the northeastern area and the upper water fishery in the southwest area), seasonal or geographic separate management regimene could be applied to the fishery. That kind of approach would also account for depth separation. This would reduce the risk of overexploitation or depletion of the possibly separate units, which would occur if they would be managed under a common TAC.

Table 10.1.1 Pelagic $S$. mentella. Landings (in tonnes) by area as used by the Working Group. Due to the lack of area reportings for some countries, the exact share in Subareas XII and XIV is just approximate in latest years.

| Year | Va | Vb | VI |  | XII |  | XIV | NAFO 1F | NAFO 2J | NAFO 2H | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0 |  | 0 | 0 |  | 39,783 | 20,798 |  |  |  | 60,581 |
| 1983 | 0 |  | 0 | 0 |  | 60,079 | 155 |  |  |  | 60,234 |
| 1984 | 0 |  | 0 | 0 |  | 60,643 | 4,189 |  |  |  | 64,832 |
| 1985 | 0 |  | 0 | 0 |  | 17,300 | 54,371 |  |  |  | 71,671 |
| 1986 | 0 |  | 0 | 0 |  | 24,131 | 80,976 |  |  |  | 105,107 |
| 1987 | 0 |  | 0 | 0 |  | 2,948 | 88,221 |  |  |  | 91,169 |
| 1988 | 0 |  | 0 | 0 |  | 9,772 | 81,647 |  |  |  | 91,419 |
| 1989 | 0 |  | 0 | 0 |  | 17,233 | 21,551 |  |  |  | 38,784 |
| 1990 | 0 |  | 0 | 0 |  | 7,039 | 24,477 | 385 |  |  | 31,901 |
| 1991 | 0 |  | 0 | 0 |  | 10,061 | 17,089 | 458 |  |  | 27,608 |
| 1992 | 1,968 |  | 0 | 0 |  | 23,249 | 40,745 |  |  |  | 65,962 |
| 1993 | 2,603 |  | 0 | 0 |  | 72,529 | 40,703 |  |  |  | 115,835 |
| 1994 | 15,472 |  | 0 | 0 |  | 94,189 | 39,028 |  |  |  | 148,689 |
| 1995 | 1,543 |  | 0 | 0 |  | 132,039 | 42,260 |  |  |  | 175,842 |
| 1996 | 4,744 |  | 0 | 0 |  | 42,603 | 132,975 |  |  |  | 180,322 |
| 1997 | 15,301 |  | 0 | 0 |  | 19,822 | 87,812 |  |  |  | 122,935 |
| 1998 | 40,612 |  | 0 | 0 |  | 22,446 | 53,910 |  |  |  | 116,968 |
| 1999 | 36,524 |  | 0 | 0 |  | 24,085 | 48,521 | 534 |  |  | 109,665 |
| 2000 | 44,677 |  | 0 | 0 |  | 19,862 | 50,722 | 10,815 |  |  | 126,076 |
| 2001 | 28,148 |  |  |  |  | 31,751 | 62,148 | 5,299 | 1,284 | 208 | 128,838 |
| $2002{ }^{1}$ | 37,388 |  |  |  |  | 23,954 | 62,684 | 7,639 |  |  | 131,665 |

${ }^{1}$ ) Provisional data

Table 10.1.2 Pelagic $S$. mentella catches (in tonnes) in ICES Div. Va, Subareas XII, XIV and NAFO Div. 1F, 2H and 2J by countries used by the Working Group.

| Year | Bulgaria | Canada | Estonia | Faroes | France | Germany ${ }^{3}$ | Greenland | Iceland | Japan | Latvia | Lithuania | Netherland | Norway | Poland | Portugal | Russia ${ }^{2}$ | Spain | UK | Ukraine | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 581 | 0 | 60,000 | 0 | 0 | 0 | 60,581 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 155 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 60,079 | 0 | 0 | 0 | 60,234 |
| 1984 | 2,961 | 0 | 0 | 0 | 0 | 989 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 239 | 0 | 60,643 | 0 | 0 | 0 | 64,832 |
| 1985 | 5,825 | 0 | 0 | 0 | 0 | 5,438 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 135 | 0 | 60,273 | 0 | 0 | 0 | 71,671 |
| 1986 | 11,385 | 0 | 0 | 5 | 0 | 8,574 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 149 | 0 | 84,994 | 0 | 0 | 0 | 105,107 |
| 1987 | 12,270 | 0 | 0 | 382 | 0 | 7,023 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 25 | 0 | 71,469 | 0 | 0 | 0 | 91,169 |
| 1988 | 8,455 | 0 | 0 | 1,090 | 0 | 16,848 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 65,026 | 0 | 0 | 0 | 91,419 |
| 1989 | 4,546 | 0 | 0 | 226 | 0 | 6,797 | 567 | 3,816 | 0 | 0 | 0 | 0 | 0 | 112 | 0 | 22,720 | 0 | 0 | 0 | 38,784 |
| 1990 | 2,690 | 0 | 0 | 0 | 0 | 7,957 | 0 | 4,537 | 0 | 0 | 0 | 0 | 7,085 | 0 | 0 | 9,632 | 0 | 0 | 0 | 31,901 |
| 1991 | 0 | 0 | 2,195 | 115 | 0 | 571 | 0 | 8,783 | 0 | 0 | 0 | 0 | 6,197 | 0 | 0 | 9,747 | 0 | 0 | 0 | 27,608 |
| 1992 | 628 | 0 | 1,810 | 3,765 | 2 | 6,447 | 9 | 15,478 | 0 | 780 | 6,656 | 0 | 14,654 | 0 | 0 | 15,733 | 0 | 0 | 0 | 65,962 |
| 1993 | 3,216 | 0 | 6,365 | 7,121 | 0 | 17,813 | 710 | 22,908 | 0 | 6,803 | 7,899 | 0 | 14,990 | 0 | 0 | 25,229 | 0 | 0 | 2,782 | 115,835 |
| 1994 | 3,600 | 0 | 17,875 | 2,896 | 606 | 17,152 | 0 | 53,332 | 0 | 13,205 | 7,404 | 0 | 7,357 | 0 | 1,887 | 17,814 | 0 | 0 | 5,561 | 148,689 |
| 1995 | 3,800 | 602 | 16,854 | 5,239 | 226 | 18,985 | 1,856 | 34,631 | 1,237 | 5,003 | 22,893 | 13 | 7,457 | 0 | 5,125 | 44,182 | 4,554 | 0 | 3,185 | 175,842 |
| 1996 | 3,500 | 650 | 7,092 | 6,271 | 0 | 21,245 | 3,537 | 62,903 | 415 | 1,084 | 10,649 | 0 | 6,842 | 0 | 2,379 | 45,748 | 7,229 | 260 | 518 | 180,322 |
| 1997 | 0 | 111 | 3,720 | 3,945 | 0 | 20,476 | 0 | 41,276 | 31 | 0 | 0 | 0 | 3,179 | 886 | 3,674 | 36,930 | 8,707 | 0 | 0 | 122,935 |
| 1998 |  |  | 3,968 | 7,474 | 0 | 18,047 | 1,463 | 48,519 | 31 |  | 1,768 |  | 1,139 | 12 | 4,133 | 25,837 | 4,577 | 0 |  | 116,968 |
| 1999 |  |  | 2,108 | 4,656 | 0 | 16,489 | 4,269 | 43,923 | 0 |  |  |  | 5,435 | 6 | 4,302 | 17,957 | 10,332 | 188 |  | 109,665 |
| 2000 |  |  | 11,811 | 2,837 | 0 | 12,499 | 4,204 | 45,232 | 0 |  | 450 |  | 5,194 | 0 | 3,731 | 29,224 | 10,894 | 0 | 0 | 126,076 |
| 2001 |  |  | 887 | 7,981 |  | 10,669 | 3,309 | 42,472 |  |  | 15,689 |  | 5,222 |  | 2,514 | 30,012 | 10,083 |  |  | 128,838 |
| $2002{ }^{1}$ |  |  |  | 4,246 |  | 13,191 | 4,264 | 44,430 |  | 1,144 | 14,656 |  | 5,451 |  | 3,164 | 32,169 | 8,950 |  |  | 131,665 |

1) Provisional data. 1991.
2) Former GDR and GFR.

Table 10.1.3 Pelagic $S$. mentella landings (in tonnes) in 2002 by countries and depth (A), and in 1997-2001 by depth (B). (Working Group figures and/or as reported to NEAFC).
$\left.\begin{array}{cccc}\hline \text { A. } & \text { Total } & \text { not splitted } & \begin{array}{c}\text { shallower than } 600 \\ \mathrm{~m}\end{array}\end{array} \begin{array}{c}\text { deeper than } \\ 600 \mathrm{~m}\end{array}\right)$

| B. | Total | not splitted | shallower than <br> 600 m | deeper than <br> 600 m |
| :---: | :---: | :---: | :---: | :---: |
| 1996 | 180,322 | $43 \%$ | $14 \%$ | $43 \%$ |
| 1997 | 122,935 | $37 \%$ | $20 \%$ | $43 \%$ |
| 1998 | 116,968 | $14 \%$ | $20 \%$ | $66 \%$ |
| 1999 | 109,665 | $22 \%$ | $14 \%$ | $64 \%$ |
| 2000 | 126,076 | $46 \%$ | $15 \%$ | $39 \%$ |
| 2001 | 128,838 | $46 \%$ | $19 \%$ | $35 \%$ |
| 2002 | 131,665 | $50 \%$ | $10 \%$ | $40 \%$ |

Table 10.1.4 Results of dividing the Icelandic pelagic redfish catch ( t ) according to the Icelandic samples from the fishery.

|  | oceanic | Deep sea | Not classified | Catch <br> Oceanic | Catch <br> Deep sea | Total <br> Catch |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1995 | $72 \%$ | $27 \%$ | $0 \%$ | 25186 | 9445 | 34631 |
| 1996 | $45 \%$ | $52 \%$ | $3 \%$ | 29182 | 33721 | 62903 |
| 1997 | $36 \%$ | $64 \%$ | $0 \%$ | 14859 | 26417 | 41276 |
| 1998 | $10 \%$ | $85 \%$ | $4 \%$ | 5504 | 46780 | 52284 |
| 1999 | $15 \%$ | $85 \%$ | $0 \%$ | 6765 | 37159 | 43924 |
| 2000 | $5 \%$ | $95 \%$ | $0 \%$ | 2455 | 42507 | 45008 |
| 2001 | $34 \%$ | $66 \%$ |  | 4423 | 27999 | 42423 |
| $2002^{*}$ | $14 \%$ | $86 \%$ |  | 6229 | 38262 | 44491 |

[^6]Table 10.2.1 Pelagic redfish $S$. mentella. Time-series of survey results, areas covered, hydro-acoustic abundance and biomass estimates shallower and deeper than 500 m (based on standardized trawl catches converted into hydro-acoustic estimates derived from linear regression models).

| Year | $\begin{gathered} \text { Area } \\ \text { covered } \\ \left(1000 \mathrm{NM}^{2}\right) \end{gathered}$ | $\begin{gathered} \text { Acoustic } \\ \text { estimates } \\ <500 \mathrm{~m}\left(10^{6}\right. \\ \text { ind.) } \end{gathered}$ | $\begin{gathered} \text { Acoustic } \\ \text { estimates } \\ <500 \mathrm{~m}(1000 \mathrm{t}) \end{gathered}$ | Trawl estimates < 500 m (10 ${ }^{6}$ ind.) | Trawl estimates < 500 m (1000 t) | Trawl estimates $\begin{gathered} >500 \mathrm{~m}\left(10^{6}\right. \\ \text { ind.) } \end{gathered}$ | $\begin{gathered} \text { Trawl } \\ \text { estimates } \\ >500 \mathrm{~m}(1000 \\ \mathrm{t}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 105 | 3498 | 2235 |  |  |  |  |
| 1992 | 190 | 3404 | 2165 |  |  |  |  |
| 1993 | 121 | 4186 | 2556 |  |  |  |  |
| 1994 | 190 | 3496 | 2190 |  |  |  |  |
| 1995 | 168 | 4091 | 2481 |  |  |  |  |
| 1996 | 253 | 2594 | 1576 |  |  |  |  |
| 1997 | 158 | 2380 | 1225 |  |  |  |  |
| 1999 | 296 | 1165 | 614 |  |  | 638 | 497 |
| 2001 | 420 | 1370 | 716 | 1955 | 1075 | 1446 | 1057 |

Table 10.2.2 Pelagic redfish S. mentella. 1999 and 2001 survey biomass estimates and area splitting between NAFO and NEAFC Convention areas by depth (shallower and deeper than 500 m ).

|  | NAFO (000 t) | NAFO \% | NEAFC (000 t) | NEAFC \% | Sum (000 t) |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 1999 shallower than $500 \mathrm{~m} *$ | 540 | 46.3 | 626 | 53.7 | 1166 |
| 1999 deeper than 500 m | 74 | 11.6 | 564 | 88.4 | 638 |
| 1999 Sum | 614 | 34.0 | 1190 | 66.0 | 1804 |
|  |  |  |  |  |  |
| 2001 shallower than 500 m | 686 | 63.8 | 390 | 36.2 | 1076 |
| 2001 deeper than 500 m | 165 | 15.6 | 892 | 84.4 | 1057 |
| 2001 Sum | 851 | 39.9 | 1282 | 60.1 | 2133 |

* acoustically measured

Table 10.2.3 Pelagic $S$. mentella. Catch per unit effort ( $\mathrm{t} / \mathrm{h}$ ) by country in Subareas XII and XIV.

| Year | Bulgaria | Germany $^{2}$ | Iceland | Norway | USSR-Russia (BMRT) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | - | - | - | - | 1.99 |
| 1983 | - | - | - | - | 1.60 |
| 1984 | 1.25 | - | - | - | 1.48 |
| 1985 | 1.85 | - | - | - | 1.68 |
| 1986 | 2.04 | - | - | - | 1.35 |
| 1987 | 1.22 | 0.79 | - | - | 1.10 |
| 1988 | 0.82 | 1.28 | - | - | 1.00 |
| 1989 | - | 0.70 | 1.11 | - | 1.00 |
| 1990 | - | 0.89 | 1.02 | 1.09 | 0.99 |
| 1991 | - | - | 1.52 | 1.42 | 0.80 |
| 1992 | - | - | 1.66 | 1.79 | 0.63 |
| 1993 | - | - | 3.27 | 2.02 | 0.63 |
| 1994 | - | - | 2.64 | 2.83 | 1.70 |
| 1995 | - | 2.06 | 2.00 | 2.05 | 1.00 |
| 1996 | - | 1.45 | 1.74 | 1.20 | 1.30 |
| 1997 | - | 1.31 | 1.11 | 0.66 | -3 |
| 1998 | - | 1.30 | 1.56 | 0.75 | - |
| 1999 | - | 0.97 | 1.55 | 0.97 | - |
| 2000 | - | 1.05 | 1.98 | 1.12 | - |
| 2001 | - | 0.91 | 1.40 | 0.88 | - |
| $2002^{1}$ | - | 1.14 | 1.90 | 1.23 | 0.89 |

[^7]3 1997-2001 Russian effort data are only available as fishing days

Table 10.2.4.a Results of the GLM model to calculate standardized CPUE for all pelagic redfish fishery, including single tow data from Germany (1995-2002), Iceland (1995-2002), Greenland (1999-2002), Faroe Island (1995-2001), Russia (1997-2001) and Norway (1995-2002). Note that the full output is not shown (afli=catch; ltogtimi=trawling time; ices = ices statistical squere; skip= vessel).

```
Analysis of Deviance Table
Quasi-likelihood model
Response: afli
Terms added sequentially (first to last)
                        Df Deviance Resid. Df Resid. Dev F Value Pr(F)
            NULL 36012 181689865
        ltogtimi 1 1404554 36011 180285311 397.0146 0
    factor(yy) 7 7548837 36004 172736474 304.8246 0
    factor(mm) 11 6036235 35993 166700238 155.1106 0
factor(skip) 84 28456849 35909 138243390 95.7581 0
factor(ices) 230 9371695 35679 128871694 11.5175 0
Call: glm(formula = afli ~ ltogtimi + factor(yy) + factor(mm) + factor(skip) +
factor(ices), family = quasi(link = log, variance = mu), data = testdata)
Deviance Residuals:
    Min 1Q Median 3Q Max
    -251.8473-42.12827-4.760622 32.18028 409.4896
```

Coefficients:

|  | Value | Std. Error | t value |
| ---: | ---: | ---: | ---: |
| (Intercept) | 8.941645120 | 0.510787055 | 17.50562202 |
| ltogtimi | 0.156644728 | 0.006553003 | 23.90426746 |
| factor(yy)1996 | 0.044604410 | 0.018704515 | 2.38468677 |
| factor(yy)1997 | -0.241500209 | 0.017338001 | -13.92895327 |
| factor(yy)1998 | -0.122250482 | 0.017869202 | -6.84140676 |
| factor (yy)1999 | -0.193460622 | 0.017892690 | -10.81227140 |
| factor (yy)2000 | -0.019415520 | 0.018040945 | -1.07619194 |
| factor (yy)2001 | -0.058639607 | 0.017632108 | -3.32572859 |
| factor (yy)2002 | 0.081719833 | 0.018402326 | 4.44073398 |
| factor $(\mathrm{mm}) 2$ | -1.433731141 | 0.447433982 | -3.20434120 |
| factor $(\mathrm{mm}) 3$ | -0.216129982 | 0.270511180 | -0.79896876 |
| factor $(\mathrm{mm}) 4$ | 0.168043623 | 0.267596974 | 0.62797281 |
| factor $(\mathrm{mm}) 5$ | 0.407939738 | 0.267589020 | 1.52450103 |
| factor $(\mathrm{mm}) 6$ | 0.307642616 | 0.267602810 | 1.14962401 |
| factor $(\mathrm{mm}) 7$ | 0.138641428 | 0.267760706 | 0.51778108 |
| factor $(\mathrm{mm}) 8$ | 0.265633370 | 0.267983037 | 0.99123203 |
| factor $(\mathrm{mm}) 9$ | 0.202804060 | 0.268085695 | 0.75648967 |
| factor $(\mathrm{mm}) 10$ | 0.100219041 | 0.268289993 | 0.37354744 |
| factor $(\mathrm{mm}) 11$ | 0.016326485 | 0.269664939 | 0.06054360 |
| factor $(\mathrm{mm}) 12$ | -0.131193760 | 0.296476477 | -0.44250985 |

(Dispersion Parameter for Quasi-likelihood family taken to be 3537.789 )
Null Deviance: 181689865 on 36012 degrees of freedom Residual Deviance: 128871694 on 35679 degrees of freedom Number of Fisher Scoring Iterations: 4

Table 10.2.4.b Results of the GLM model to calculate standardized CPUE for pelagic redfish fishery, by depths shallower than 500 m (south-western area) including singlel tow data from Germany (1995-2002), Iceland (1995-2002), Greenland (1999-2002), Faroe Island (1995-2001), Russia (1997-2001) and Norway (1995-2002). Note that the full output is not shown.

```
Analysis of Deviance Table
Quasi-likelihood model
Response: afli
Terms added sequentially (first to last)
                            Df Deviance Resid. Df Resid. Dev F Value Pr(F)
                            NULL 9213 52614579
        ltogtimi 1 316778 9212 52297802 87.0662 0
        factor(yy) 7 4303329 9205 47994472 168.9670 0
        factor(mm) 9 3437250 9196 44557223 104.9697 0
factor(skip) 67 8856328 9129 35700895 36.3307 0
factor(ices) 131 2557084 8998 33143811 5.3650 0
Call: glm(formula = afli ~ ltogtimi + factor(yy) + factor(mm) + factor(skip) +
factor(ices), family = quasi(link = log, variance = mu), data = testdata)
Deviance Residuals:
    Min 1Q Median 3Q Max
    -239.9342 -39.58192 -5.26949 31.50438 373.4299
```

Coefficients:

|  | Value | Std. Error | t value |
| ---: | ---: | ---: | ---: |
| (Intercept) | 8.04681561 | 0.52609392 | 15.2953974 |
| ltogtimi | 0.18135630 | 0.01207254 | 15.0222096 |
| factor (yy)1996 | 0.07412931 | 0.05551622 | 1.3352730 |
| factor (yy)1997 | -0.15569141 | 0.04114485 | -3.7839825 |
| factor (yy)1998 | -0.07993046 | 0.03846676 | -2.0779101 |
| factor (yy)1999 | -0.43158288 | 0.03580414 | -12.0539925 |
| factor (yy)2000 | -0.12360495 | 0.03800673 | -3.2521858 |
| factor (yy)2001 | -0.05944216 | 0.03598037 | -1.6520722 |
| factor $(y y) 2002$ | -0.07248747 | 0.03899156 | -1.8590554 |
| factor $(\mathrm{mm}) 4$ | 1.52691065 | 0.25476278 | 5.9934605 |
| factor $(\mathrm{mm}) 5$ | 1.73895938 | 0.24385059 | 7.1312493 |
| factor $(\mathrm{mm}) 6$ | 1.24732778 | 0.24404510 | 5.1110544 |
| factor $(\mathrm{mm}) 7$ | 1.16946407 | 0.24290056 | 4.8145796 |
| factor $(\mathrm{mm}) 8$ | 1.41474266 | 0.24220915 | 5.8409960 |
| factor $(\mathrm{mm}) 9$ | 1.32889393 | 0.24214161 | 5.4880858 |
| factor $(\mathrm{mm}) 10$ | 1.23371829 | 0.24256991 | 5.0860317 |
| factor $(\mathrm{mm}) 11$ | 0.90064532 | 0.24630319 | 3.6566531 |
| factor $(\mathrm{mm}) 12$ | 1.22066784 | 0.28604795 | 4.2673540 |

Table 10.2.4.c Results of the GLM model to calculate standardized CPUE for pelagic redfish fishery, by depths deeper than 500 m (south-western area) including singlel tow data from Germany (1995-2002), Iceland (19952002), Greenland (1999-2002), Faroe Island (1995-2001), Russia (1997-2001) and Norway (19952002). Note that the full output is not shown.

```
Analysis of Deviance Table
Quasi-likelihood model
Response: afli
Terms added sequentially (first to last)
                Df Deviance Resid. Df Resid. Dev F Value Pr(F)
            NULL 26784 128997336
```



```
    factor(yy) 7 8317519 26776 119567371 361.8469 0
    factor(mm) 11 3461155 26765 116106216 95.8203 0
factor(skip) 79 21486982 26686 94619234 82.8280 0
factor(ices) 98 4603244 26588 90015990 14.3043 0
Call: glm(formula = afli ~ ltogtimi + factor(yy) + factor(mm) + factor(skip) +
factor(ices), family = quasi(link = log, variance = mu), data = testdata)
Deviance Residuals:
    Min 1Q Median 3Q Max
-227.6338-41.14726-4.220292 32.0225 347.7267
```

Coefficients:

|  | Value | Std. Error | t value |
| ---: | ---: | ---: | ---: |
| (Intercept) | 8.434173631 | 0.293986897 | 28.6889440 |
| ltogtimi | 0.164426840 | 0.007864836 | 20.9065829 |
| factor (yy)1996 | 0.006160440 | 0.022212893 | 0.2773363 |
| factor (yy)1997 | -0.308500874 | 0.021407430 | -14.4109256 |
| factor (yy)1998 | -0.178554522 | 0.022136918 | -8.0659160 |
| factor (yy)1999 | -0.142140342 | 0.022462434 | -6.3279136 |
| factor (yy)2000 | -0.048003716 | 0.022319939 | -2.1507100 |
| factor $(\mathrm{yy}) 2001$ | -0.131483338 | 0.022143955 | -5.9376629 |
| factor $(\mathrm{yy}) 2002$ | 0.061631469 | 0.022672005 | 2.7183951 |
| factor $(\mathrm{mm}) 2$ | -1.516024853 | 0.431386362 | -3.5143087 |
| factor $(\mathrm{mm}) 3$ | -0.246928657 | 0.260973190 | -0.9461840 |
| factor $(\mathrm{mm}) 4$ | 0.117444720 | 0.257913521 | 0.4553647 |
| factor $(\mathrm{mm}) 5$ | 0.353140305 | 0.257910067 | 1.3692382 |
| factor $(\mathrm{mm}) 6$ | 0.278962868 | 0.257922184 | 1.0815776 |
| factor $(\mathrm{mm}) 7$ | 0.151546367 | 0.258098647 | 0.5871645 |
| factor $(\mathrm{mm}) 8$ | 0.078455882 | 0.259091372 | 0.3028116 |
| factor $(\mathrm{mm}) 9$ | -0.029399779 | 0.260322807 | -0.1129359 |
| factor $(\mathrm{mm}) 10$ | -0.167667805 | 0.261546422 | -0.6410633 |
| factor $(\mathrm{mm}) 11$ | 0.127224735 | 0.262556194 | 0.4845619 |
| factor $(\mathrm{mm}) 12$ | -0.865802042 | 0.360270199 | -2.4032019 |



Figure 10.1.1 Depth distribution of Icelandic trawl hauls for oceanic redfish as reported in the log-books since Iceland began its oceanic redfish fishery in 1989. X -axis $=$ day of year; Y axis $=$ depth.


Figure 10.1.2 Length distributions from landings of pelagic S. mentella in 1995-2002.


Figure 10.1.3 Length distributions from German and Spanish landings of pelagic S. mentella in 1995-2002, divided by depths shallower and deeper than 500 m .


Figure 10.2.1 Pelagic redfish S. mentella. Survey catches in June/July 2001 shallower than 500 m depth (black) and deeper than 500 m depth (grey).


Figure 10.2.3.a Trends in CPUE of pelagic $S$. mentella fishery in the Irminger Sea, shallower than 500 m , and estimated acoustic biomass from surveys.


Figure 10.2.3.b Trends in CPUE of pelagic $S$. mentella fishery in the Irminger Sea, deeper than 500 m , and estimated trawl biomass from surveys.


Figure 10.2.3.c
Standardised CPUE, as calculated by using data from Germany (1995-2002), Iceland (1995-2002), Greenland (1999-2002), Faroe Island (1995-2001), Russia (1997-2001) and Norway (1995-2002) in the GLM model (see chapter 10.2.2.), divided by depths shallower (south-western area) and deeper than 500 m (northeastern area) and both depth layers (areas) combined (All data. 95\% confidence limits are shown. Further details of the GLM models are given in Table 10.2.4

The following is a list of working documents that were made available both before and during the WG meeting.

WD 01: Hans-Joachim Rätz, Jens Ulleweit and Kai Panten. Data on German catches and effort for Greenland halibut (Reinhardtius hippoglossoides), demersal redfish (Sebastes marinus and deep sea S. mentella), and Atlantic cod (Gadus morhua) in ICES Div. Va, Vb, VIa and XIV, 1995-2002:

WD 02: Hans-Joachim Rätz, Thorsteinn Sigurðsson and Christoph Stransky. Abundance and length composition for Sebastes marinus L., deep sea S. mentella and juvenile redfish (Sebastes spp.) off Greenland and Iceland based on groundfish surveys 1985-2002:

WD 03: Hans-Joachim Rätz. Groundfish Survey Results for Cod off Greenland (offshore component):

WD 04: Hans-Joachim Rätz, Jens Ulleweit and Kay Panten. On the German Fishery and Biological Characteristics of Oceanic Redfish (Sebastes mentella Travin):

WD 05: O.A. Jørgensen. Survey for Greenland halibut in ICES Division 14B, June-July 2002.:

WD 06: Fernando González. Report of the fishing activity of the Spanish fleet in Sebastes mentella fishery in 2002:

WD 07: S.P. Melnikov. Preliminar information about Russian fishery for the oceanic $S$. mentella in ICES subareas XII, XIV, in NAFO division 1F in 2002 and biological sampling from commercial catches.:

WD 08: Shibanov V. and S.Melnikov. Pelagic Sebastes mentella stock structure in ICES Subareas XII, XIV and NAFO Conventional Area by the results of Russian investigations:

WD 09: S. P. Melnikov, Yu. I. Bakay, I. V. Bakay, G. G. Novikov and A. N. Stroganov. Ecological and biological characteristics of redfish Sebastes mentella in Va and XIVb Divisions of ICES:

WD 10: Christoph Stransky. Shape analysis and microchemistry of redfish otoliths: investigation of geographical differences in the North Atlantic:

WD 11: Christoph Stransky, Sif Guðmundsdóttir, Porsteinn Sigurðsson, Svend Lemvig and Kjell Nedreaas. Age readings of Sebastes marinus and $S$. mentella otoliths: bias and precision between readers:

WD 12: R. Alpoim, J. Vargas and E. Santos. NEAFC Portuguese research report for 2002:

WD 13: Fernando González. Report of NAFO scientific council in 2002:

WD 14: Petur Steingrund. Correction of maturity stages in the Faroese spring groundfish survey:
WD 15: J. Boje. The fishery for Greenland halibut in ICES Div. XIVb in 20The fishery for Greenland halibut in ICES Div. XIVb in 200202:

WD 16: Marie Storr-Paulsen. Cod stock off West Greenland:

WD 17: Thorsteinn Sigurðsson. Sebastes marinus in ICES division VA. Figures and tables:

WD 18: Thorsteinn Sigurdsson, Hajo Rätz, Kjell Nedreaas, Sergei P. Melnikov and Jákup Reinert . Fishery on pelagic redfish (S.mentella, Travin):Information based on log-book data from Faroe Island, Germany, Greenland, Iceland, Norway and Russia.

WD 19: Thorsteinn Sigurðsson. Information on the Icelandic Fishery of Oceanic Redfish (S.Mentella Travin); Information based on log-book data and sampling from the commercial fFishery:

WD 20: Jákup Reinert. Some information on the Faroese redfish fishery:

WD 21: Jákup Reinert. Preliminary assessment of Faroe Haddock:
WD 22: Agnes C. Gundersen and Åge Høines. Norwegian fishery for Greenland halibut and S.marinus and in ICES Subareas XII and XIV, and pelagic Sebastes mentella in the Irminger Sea, 2001-2002:

WD 23: Thorsteinn Sigurdsson and Kristjan Kristinsson. Information on the shelf deep-sea redfish (Sebastes mentella) fishery in Division Va 2003:

WD 24: Lise Helen Ofstad. Preliminary Assessment of Faroe Saithe 2002:

WD 25: Petur Steingrund. Preliminary assessments of Faroe Plateau cod:
WD 26: The Icelandic Haddock Council. Icelandic haddock:
WD 27: Björn Ævarr Steinarsson, Einar Hjörleifsson, Höskuldur Björnsson, Ólafur Karvel Pálsson and Sigfús Alexander Schopka. External experts: Árni Magnússon, Guðmundur Guðmundsson and Porvaldur Gunnlaugsson. Icelandic cod in division Va:

WD 28: Einar Hjörleifsson. Excel Statistical Catch @ Age Model (EXCAM) analysis of Fareo haddock
WD 29: Einar Hjörleifsson. Excel Statistical Catch @ Age Model (EXCAM) analysis of Faroe cod:
WD 30: Kristján Kristinsson and Porsteinn Sigurðsson. Request from Northeast Atlantic Fisheries Commission Regarding Redfish Stocks - Terms of Reference C.:

WD 31: Arni Magnusson. 2003 Coleraine assessment of the Icelandic cod stock:
WD 32: Einar Hjörleifsson. Excel Statistical Catch @ Age Model (EXCAM) analysis of Greenland cod
WD 33: Withdrawn

WD 34: Höskuldur Björnsson. Investigation of the relationship between hidden mortality of Icelandic haddock and fishing mortalility.


[^0]:    * Preliminary
    ${ }^{1)}$ Included in Vb2.
    ${ }^{2)}$ Reported as Vb.

[^1]:    Weights in kilograms

[^2]:    *) Preliminary.

    1) Includes Vb1
    2) Included in Vb1
[^3]:    Weights in kilograms

[^4]:    ${ }^{1}$ ) Figure reflects Dohrn Bank area only due to reduced survey area.
    ${ }^{2}$ ) No estimate available for the Dohrn Bank-East Greenland area due to reduced survey area.

[^5]:    ${ }^{1}$ ) Provisional data reported by Greenland authorities
    ${ }^{2}$ ) Includes 3,000 t reported to be caught in ICES Subarea XIV
    ${ }^{3}$ ) Includes 2,741 t reported to be caught in ICES Subarea XIV
    ${ }^{4}$ ) Includes 29,513 t caught inshore

[^6]:    *Preliminary

[^7]:    2 1987-1990 reported as GDR (FVSIV)

