## REPORT OF THE

# NORTHERN PELAGIC AND BLUE WHITING FISHERIES WORKING GROUP 

ICES Headquarters<br>29 April-8 May 2003

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International Council for the Exploration of the Sea
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## TECHNICAL MINUTES

# Subgroup Northern Pelagic and Blue Whiting Fisheries Working Group (WGNPBW) 

ACFM Meeting May 2003

Thursday 30/5/2003: 9:00-18:00 Beverton room<br>Subgroup chair: Martin Pastoors<br>Presenter: Asta Gudmundsdottir<br>Reviewers: Tomas Gröhsler, Vladimir Shebanov<br>Participants: Dankert Skagen., Bengt Sjöstrand., Einar Hjörleifsson,<br>Poul Degnbol, Geoge Kornilovs<br>Observer: Ken Patterson

## GENERAL COMMENTS

The structure of the report has improved, but still substantial further improvements could be made. The report would benefit from an introductory section that links the terms of reference to the sections where they are dealt with. The tables and figures are sometimes organized rather haphazardly which makes the job of reviewing the report difficult.

For each stock the whole assessment input data and output results should be presented both in the report and the ICES server. The subgroup suggests the following sequence of tables as a minimum requirement for assessment documentation:

Input data (tables):

- catch in numbers
- catch/stock weight-at-age
- natural mortality
- maturity
- survey data
- tagging data
- larval data

Input (figures):

- catch in numbers (log scale?)
- survey indices
- larval indices


## Output (tables):

- stock numbers-at-age
- fishing mortality-at-age
- SSB at age and combined
- Stock summary
- parameter estimates
- residuals (log catchability residuals)

Output (figures):

- stock summary
- residual plots
- retrospective analysis
- uncertainty analysis (bootstrap)
- stock recruitment stuff
- biological reference points

Not all the ToR have been explicitly addresses. The special requests by Russia and coastal states have not been addressed by the WG. For WHB it just referred to Methods WG which had suggested a number of analysis to the WG. The ToR on exploitation of juvenile blue whiting was addressed in section 6.8.1 but this was not referred to. The ToR j deficiencies in the assessments was not addressed. Also the ToR k comparison with last year was not addressed.

The inability of the group to reach consensus on the Russian and Norwegian inputs to the group severely compromised ACFM position with regards to providing advice. It is then not enough to state that joint calculations should be undertaken (e.g. p. 7) but also how this will be achieved. The WG should draw conclusions from different assessments.

Methods description: a short description of the salient features of the methods is required. This should highlight the main features of the model and which specific problem it intends to solve. The subgroup reiterates that it is not sufficient to refer to a website for documentation of a method, because it should at least be possible to obtain a printed copy of the description/manual somewhere.

Colors are useless in printed document (e.g. figure 3.3.1).
Is it useful to have information on sampling level (No. sampled, No. measured etc. ) for the commercial landings in each stock section. The subgroup suggests to include all the output from the Salloc program in the report.

It would be useful to include a map in the introduction section showing the distribution areas of the different stocks.

## Ecological considerations

The section on ecological and environmental information was informative and well written. However, it was noted that as already addressed last year that there was hardly any linkage with the stock chapters to the chapter on ecological considerations. The section as a whole would benefit with indicating specific linkages to other sections.

A section on phytoplankton in the Barents Sea is missing.

What conclusion can be drawn from section 2.6.

## Norwegian spring spawning herring (HER-NOSS)

The structure of this section of the report has improved compared to last year.

It is unacceptable that information pertinent to the assessment is kept at a website in Norway rather than on the ICES server. The input and output of the assessment runs could not be found on the ICES server. The subgroup was very critical about the procedures that have been followed to assess this stock and stresses that explaining the procedures and results is just as important as obtaining the results in the first place. The people responsible for this assessment are requested to adhere to the basic list of input and output requirements which is presented in section 1 of the technical minutes.

SeaStar uses tuning on the largest year classes only. How is the decision reached when a year class is considered to be strong. The choice seems to be arbitrary. Could this be illustrated using e.g. the catch data?

The SeaStar assessment has the fundamental assumption that no age dependence or abundance dependence in catchability of surveys applies. Can this be substantiated and has the sensivity of the model to this assumption been tested?

The group describes the SeaStar method as giving an objective way of choosing weights on different sources of information. Nevertheless many arbitrary choices are made in this assessment: to leave out data from the assessment: larval data, tagging data, etc. This violates the objectivity. The WG is requested to provide better analysis/explanations of the reasons for leaving out observations to avoid the impression of arbitrariness.

The subgroup recommended rather than to model the ageing errors directly to use a plus group at a lower age.

It is necessary to present any assessment with all data in and with settings similar to last year in order to be able to make comparisons. After that, explorations can be made with different model settings or data configurations.

The final SeaStar assessment was found to incorporate inconsistencies: mean weights, maturities, SSB estimates. During the subgroup a new run was set up and the results are incorporated in appendix 2 . The new run does not provide major changes in the perception of the stock.

ISVPA Gives very different results from the SeaStar model. Youngest ages in catch data are difficult to follow. Very high estimates of juveniles (e.g. 5 year olds). Residuals dependent on year class (either all positive or all negative). Could this model also be run with less constraints on the residual matrix, so that the strong year class effects in the residuals would disappear, as the ISVPA is now not longer based on catch data only?

The bootstrap results for ISVPA are very difficult to explain as the final ISVPA assessment is towards the outer bounds of the confidence interval. The subgroup wonders whether there is a programming error that causes this result or that the constraints in the model are too tight.

The subgroup has selected the SeaStar assessment as the basis for advice. The reasons for this choice are:

- Youngest ages in catch data are difficult to follow (i.e. when they enter the fishery).
- ISVPA has a high SoP difference between the predicted and observed catches in a recent year
- ISVPA residuals seem to be very dependent on year class (either all positive or all negative).

Nevertheless, the subgroup has recommended an external review be conducted in order to produce a accepted assessment method for this stock.

Section 3.4.4 on short-term projections is too short. A description is required on the basis for the input and comments on the results. The selection pattern used for the SeaStar prediction seems to be inappropriate as may substantially overestimate the selection on the youngest ages. The ISVPA selection pattern seems to behave better even though it looks likes it has been forced to a function

Medium analysis were not carried out during the working group. There were medium-term analysis found on the IMR website, but there was no documentation on the basis for the analysis. The subgroup decided not to incorporate the results in the advice.

## Barents Sea Capelin ()

The subgroup noticed that this stock is left in between the may and october ACFM meetings and concluded that no reviews were presented for recent years. The group reviewed the assessment during it's discussion but did not provide specific comments.

## Capelin in the Iceland-East, Greenland, Jan Mayen area (CAP-ICEL)

Define the term fishable stock in the beginning of the text. It is also questionable if the tables could be fewer.

## Blue whiting (WHB-COMB)

Structure of the section has improved and is now very clear.
All assessment input data and output results of the final assessment should be presented (including parameter estimates and set up of the assessment model).

The subgroup suggested to have a further look at assembling survey data from national basis that is collected for other purposes (e.g. all the herring surveys). Refer to Planning group on Norwegian Sea.

AMCI: The plus group assumptions/differences should be described/explained.

AMCI with Acoustic surveys as SSB index was dismissed because it did not confirm to the catch ratio analysis. This is a way of tuning the assessment to prior knowledge which can hardly be defended. If you already know the answer, why would you need to do an assessment?

Final assessment with AMCI : Why changing age of recruitment from 0.5 in 2001 to 1 in 2002?

The subgroup has selected the AMCI assessment as the basis for advice. The reasons for this choice are:

- Similar trends in fishing mortality from the catch and survey data
- ISVPA residuals seem to be very dependent on year class (either all positive or all negative).


## Icelandic summer spawning herring (HER-VASU)

Information on sampling is missing.

The ageing problem can possibly be overcome by using a plus group at a younger age.

All the input data should be presented: i.e. Natural mortality (NATMOR), proportion F before spawning (FPROP), proportion M before spawning (MPROP).

What is the general definition of age (ie. Age 5 (Age 6 on 1 January))

## Appendix 1: detailed comments

## Section 2

Section 2.3.2 Predation by cod: Why presenting two conflicting results.
Section 2.4.5: $\mathrm{R}^{2}=0.66$ in the text is not in line with the value presented in Figure 2.4.5.3 of 0.5302

## Section 3

Is the prognosis of herring condition and weight-at-age only used in short-term forecast using SeaStar?
Numbers in Figure 3.2.3.5 are hardly readable
SeaStar Manual has no description of input and output formats.
Colour problem in Figure 3.3.3
Column in Table 3.3.4.1 not readable!

## Section 5

Units in Table 5.3.1.1 are missing.
Labelling of column in Table 5.3.2.1.2 is wrong.
Section 5.3.2.2 : 61.5 billions not corresponding with 62.5 billions in Table 5.3.2.3.
Figures 5.5.3.1 \& 5.5.3.2 are missing

## Section 6

Table 6.3.4.1 is missing.
Section 6.3.5 'The value of the natural mortality M for blue whiting' could be shortened to 'Natural Mortality'.
Tables 6.4.1-6.4.3 (2003 is missing in 6.4.2.-6.4.3) are duplicated in Tables 6.4.1.1-6.4.1.3 (which are including less information). Why using different age spans (1-9+ and 1-11+)?
Lines in Figure 6.4.4.1.2 are not readable? To be comparable to results given in Figure 6.4.4.1.1: Why not using the same time periods ( $<91,91-94$ and 95-97)?
ICA run compered to 2002: Why changing the years of separable constraint? Age 1 was downweighted in the AMCI final run. Is there any reason to not downweighting age 1 in ICA.
Units in Table 6.4.5.2.1-6.4.5.2.3 are missing
Section 6.4.5.3 Comparison: Why not including the results from ICA?
Tables 6.5.1 and 6.5.2 are identical.
Section on blue whiting: wrong reference in the text: ICES 2003/ACFM :15, it should be replaced by ICES 2003/D:03?

## Section 7

Table 7.1.2 Catch in numbers: This table should go into section 7.2
It would be nice to present a figure showing the age composition?!
Table 7.3.1: Explanation of large year class 2000 is missing (not only 1999!)
Chapter 7.4.1 Adapt VPA: Not referring to Table 7.4.1-7.4.3 in the text! Units in Tables 7.4.2 and 7.4.3 are missing! What is the meaning of WF 5-15?

Appendix 2: input and output of final SeaStar assessment on Norwegian Spring Spawing Herring.
Table 1. Catch in numbers used by SeaStar.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 5112600. | 2000000. | 600000. | 276200. | 184800. | 185500. | 547000. | 628600. | 79500. | 88600. | 109500. | 86900. | 194500. | 368300. | 66400. | 107000. | 237300. |
| 1951 | 1635500. | 7607690. | 400000. | 1636600. | 383800. | 172400. | 164400. | 515600. | 602000. | 77100. | 82700. | 103100. | 107600. | 253500. | 348000. | 47400. | 305100. |
| 1952 | 13721600. | 9149700. | 1232900. | 39300. | 60500. | 602300. | 136300. | 204500. | 380200. | 377900. | 79200. | 85700. | 107700. | 106800. | 186500. | 256300. | 308100. |
| 1953 | 5697200. | 5055000. | 581300. | 740100. | 46600. | 100900. | 355600. | 81900. | 110900. | 314100. | 394900. | 61700. | 91200. | 94100. | 98800. | 215500. | 514900 |
| 1954 | 10675990. | 7071090. | 855400. | 266300. | 1435500. | 142900. | 236000. | 490300. | 128100. | 199800. | 440400. | 460700. | 88400. | 100600. | 133000. | 126800. | 676400. |
| 1955 | 5175600. | 2871100. | 510100. | 93000. | 276400. | 2045100. | 114300. | 189600. | 274700. | 85300. | 193400. | 295600. | 203200. | 58700. | 84600. | 103600. | 477000. |
| 1956 | 5363900. | 2023700. | 627100. | 116500. | 251600. | 314200. | 2555100. | 110000. | 203900. | 264200. | 130700. | 198300. | 272800. | 163300. | 63000. | 88900. | 476200. |
| 1957 | 5001900. | 3290800. | 219500. | 23300. | 373300. | 153800. | 228500. | 1985300. | 72000. | 127300. | 182500. | 88400. | 121200. | 149300. | 131600. | 337000 | 24770 |
| 1958 | 9666990. | 2798100. | 666400. | 17500. | 17900. | 110900. | 89300. | 194400. | 973500. | 70700. | 123000. | 200900. | 98700. | 77400. | 70900. | 69400. | 186200 |
| 1959 | 17896280. | 198530. | 325500. | 15100. | 26800. | 25900. | 146600. | 114800. | 240700. | 1103800. | 88600. | 124300. | 198000. | 88500. | 77400. | 85200. | 150700. |
| 1960 | 12884310. | 13580790. | 392500. | 121700. | 18200. | 28100. | 24400. | 96200. | 73300. | 203900. | 1163000. | 85200. | 129700. | 153500. | 56700. | 47200. | 121700. |
| 1961 | 6207500. | 16075600. | 2884800. | 31200. | 8100. | 4100. | 15000. | 19400. | 61600. | 49200. | 136100. | 728100. | 49700. | 45000. | 63000. | 21700. | 38400. |
| 1962 | 3693200. | 4081100. | 1041300. | 1843800. | 8000. | 3100. | 7200. | 20200. | 11900. | 59100. | 52600. | 117000. | 813500. | 44200. | 54700. | 65600. | 86700. |
| 1963 | 4807000. | 2119200. | 2045300. | 760400. | 835800. | 5300. | 1800. | 3600. | 18300. | 9300. | 107700. | 92500. | 174100. | 923700. | 79600. | 60400. | 124900. |
| 1964 | 3613000. | 2728300. | 220300. | 114600. | 399000. | 2045800. | 13700. | 1500. | 3000. | 24900. | 29300. | 95600. | 82400. | 153000. | 772800. | 45800. | 291000. |
| 1965 | 2303000. | 3780900. | 2853600. | 89900. | 256200. | 571100. | 2199700. | 19500. | 14900. | 7400. | 19100. | 40000. | 100500. | 107800. | 138700. | 704000. | 179100. |
| 1966 | 3926500. | 662800. | 1678000. | 2048700. | 26900. | 466600. | 1306000. | 2884500. | 37900. | 14300. | 17400. | 26200. | 11000. | 69100. | 72100. | 96700. | 460000. |
| 1967 | 426800. | 9877100. | 70400. | 1392300. | 3254000. | 26600. | 421300. | 1132000. | 1720800. | 8900. | 5700. | 3500. | 8500. | 8900. | 17500. | 14300. | 90100. |
| 1968 | 1783600. | 437000. | 388300. | 99100. | 1880500. | 1387400. | 14200. | 94000. | 134100. | 345100. | 2000. | 1100. | 800. | 2500. | 2600. | 1800. | 15200. |
| 1969 | 561200. | 507100. | 141900. | 188200. | 800. | 8800. | 4700. | 700. | 11700. | 33600. | 36000. | 300. | 200. | 200. | 200. | 400. | 2000. |
| 1970 | 119300. | 529400. | 33200. | 6300. | 18600. | 600. | 3300. | 3300. | 1000. | 13400. | 26200. | 28100. | 300. | 100. | 200. | 100. | 1900. |
| 1971 | 30500. | 42900. | 85100. | 1820. | 1020. | 1240. | 360. | 1110. | 1130. | 360. | 4410. | 6910. | 5450. | 1. | 20. | 120. | 0. |
| 1972 | 347100. | 41000. | 20400. | 35376. | 3476. | 3583. | 2481. | 694. | 1486. | 198. | 1. | 494. | 593. | 593. | 1. | 1. |  |
| 1973 | 29300. | 3500. | 1700. | 2389. | 25200. | 651. | 1506. | 278. | 178. | 1. | 1. | 1. | 1. | 1. | 180. | 1. |  |
| 1974 | 65900. | 7800. | 3900. | 100. | 241. | 24505. | 257. | 196. | 1. | 1. | 1. | 1. | 1. | 1. | 1. | 1. |  |
| 1975 | 30600. | 3600. | 1800. | 3268. | 132. | 910. | 30667. | 5. | 2. | 1. | 1. | 1. | 1. | 1. | 1. | 1. |  |
| 1976 | 20100. | 2400. | 1200. | 23248. | 5436. | 1. | 1. | 13086. | 1. | 1. | 1. | 1. | 1. | 1. | 1. | 1. |  |
| 1977 | 43000. | 6200. | 3100. | 22103. | 23595. | 336. | 1. | 419. | 10766. | 1. | 1. | 1. | 1. | 1. | 1. | 1. |  |
| 1978 | 20100. | 2400. | 1200. | 3019. | 12164. | 20315. | 870. | 1. | 620. | 5027. | 1. | 1. | 1. | 1. | 1. | 1. |  |
| 1979 | 32600. | 3800. | 1900. | 6352. | 1866. | 6865. | 11216. | 326. | 1. | 1. | 2534. | 1. | 1. | 1. | 1. | 1. |  |
| 1980 | 6900. | 800. | 400. | 6407. | 5814. | 2278. | 8165. | 15838. | 441. | 8. | 1. | 2688. | 1. | 1. | 1. | 1. |  |
| 1981 | 8300. | 1100. | 11900. | 4166. | 4591. | 8596. | 2200. | 4512. | 8280. | 345. | 103. | 114. | 964. | 1. | 1. | 1. |  |
| 1982 | 22600. | 1100. | 200. | 13817. | 7892. | 4507. | 6258. | 1960. | 5075. | 6047. | 121. | 37. | 37. | 121. | 1. | 1. |  |
| 1983 | 127000. | 4680. | 1670. | 3183. | 21191. | 9521. | 6181. | 6823. | 1293. | 4598. | 7329. | 143. | 40. | 143. | 860. | 1. |  |
| 1984 | 33860. | 1700. | 2490. | 4483. | 5388. | 61543. | 18202. | 12638. | 15608. | 7215. | 16338. | 6478. | 1. | 1. | 1. | 1650. | 0 |
| 1985 | 28570. | 13150. | 207220. | 21500. | 15500. | 16500. | 130000. | 59000. | 55000. | 63000. | 10000. | 31000. | 50000. | 1. | 1. | 1. | 2640. |
| 1986 | 13810. | 1380. | 3090. | 539785. | 17594. | 14500. | 15500. | 105000. | 75000. | 42000. | 77000. | 19469. | 66000. | 80000. | 1. | 1. | 2470. |
| 1987 | 13850. | 6330. | 35770. | 19776. | 501393. | 18672. | 3502. | 7058. | 28000. | 12000. | 9500. | 4500. | 7834. | 6500. | 7000. | 450. | 0. |
| 1988 | 15490. | 2790. | 9110. | 62923. | 25059. | 550367. | 9452. | 3679. | 5964. | 14583. | 8872. | 2818. | 3356. | 2682. | 1560. | 540. |  |
| 1989 | 7120. | 1930. | 25200. | 2890. | 3623. | 5650. | 324290. | 3469. | 800. | 679. | 3297. | 1375. | 679. | 321. | 260. | 1. | 0. |
| 1990 | 1020. | 400. | 15540. | 18633. | 2658. | 11875. | 10854. | 226280. | 1289. | 1519. | 2036. | 2415. | 646. | 179. | 590. | 170. | 310. |
| 1991 | 100. | 3370. | 3330. | 8438. | 2780. | 1410. | 14698. | 8867. | 218851. | 2499. | 461. | 87. | 690. | 103. | 260. | 530. | 10. |
| 1992 | 1630. | 150. | 1340. | 12586. | 33100. | 4980. | 1193. | 11981. | 5748. | 225677. | 2483. | 639. | 247. | 1236. | 1. | 1. | 0 |
| 1993 | 6570. | 130. | 7240. | 28408. | 106866. | 87269. | 8625. | 3648. | 29603. | 18631. | 410110. | 1. | 1. | 1. | 1. | 1. | 0 |
| 1994 | 430. | 20. | 8100. | 32500. | 110090. | 363920. | 164800. | 15580. | 8140. | 37330. | 35660. | 645410. | 2830. | 460. | 100. | 2070. | 0. |
| 1995 | 1. | 1. | 1130. | 57590. | 346460. | 622810. | 637840. | 231090. | 15510. | 15850. | 69750. | 83740. | 911880. | 4070. | 250. | 1. | 450. |
| 1996 | 1. | 1. | 30140. | 34360. | 713620. | 1571000. | 940580. | 406280. | 103410. | 5680. | 7370. | 66090. | 17570. | 836550. | 1. | 1. | 0 |
| 1997 | 1. | 1. | 21820. | 130450. | 270950. | 1795780. | 1993620. | 761210. | 326490. | 60870. | 20020. | 32400. | 90520. | 19120. | 370330. | 300. | 0 |
| 1998 | 1. | 1. | 82891. | 70323. | 242365. | 368310. | 1760319. | 1263750. | 381482. | 129971. | 42502. | 25343. | 3478. | 112604. | 5633. | 108514. | 1. |
| 1999 | 1. | 1. | 5029. | 137626. | 35820. | 134813. | 429433. | 1604959. | 1164263. | 291394. | 106005. | 14524. | 40040. | 7202. | 88598. | 1. | 63983. |
| 2000 | 1. | 1000. | 14395. | 84016. | 560379. | 34933. | 110719. | 404460. | 1299253. | 1045001. | 216980. | 71589. | 16260. | 22701. | 23321. | 4724. | 67087. |
| 2001 | 1. | 1. | 2076. | 102293. | 160678. | 426822. | 38749. | 95991. | 296460. | 839136. | 507106. | 73673. | 23722. | 3505. | 3356. | 1. | 22164. |
| 2002 | 1. | 1. | 61874. | 198081. | 640322. | 253823. | 322791. | 29769. | 93025. | 263090. | 658268. | 336916. | 52508. | 12335. | 6937. | 300. | 974 |

# Table 2. Stock weights-at-age used in SeaStar vpa. 

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 1. | 8. | 47. | 100. | 204. | 230. | 255. | 275. | 290. | 305. | 315. | 325. | 330. | 340. | 345. | 362. | 365. |
| 1951 | 1. | 8. | 47. | 100. | 204. | 230. | 255. | 275. | 290. | 305. | 315. | 325. | 330. | 340. | 345. | 362. | 365. |
| 1952 | 1. | 8. | 47. | 100. | 204. | 230. | 255. | 275. | 290. | 305. | 315. | 325. | 330. | 340. | 345. | 362. | 365. |
| 1953 | 1. | 8. | 47. | 100. | 204. | 230. | 255. | 275. | 290. | 305. | 315. | 325. | 330. | 340. | 345. | 362. | 365. |
| 1954 | 1. | 8. | 47. | 100. | 204. | 230. | 255. | 275. | 290. | 305. | 315. | 325. | 330. | 340. | 345. | 362. | 365. |
| 1955 | 1. | 8. | 47. | 100. | 195. | 213. | 260. | 275. | 290. | 305. | 315. | 325. | 330. | 340. | 345. | 362. | 365. |
| 1956 | 1. | 8. | 47. | 100. | 205. | 230. | 249. | 275. | 290. | 305. | 315. | 325. | 330. | 340. | 345. | 362. | 365. |
| 1957 | 1. | 8. | 47. | 100. | 136. | 228. | 255. | 262. | 290. | 305. | 315. | 325. | 330. | 340. | 345. | 362. | 365. |
| 1958 | 1. | 8. | 47. | 100. | 204. | 242. | 292. | 295. | 293. | 305. | 315. | 330. | 340. | 345. | 352. | 360. | 365. |
| 1959 | 1. | 8. | 47. | 100. | 204. | 252. | 260. | 290. | 300. | 305. | 315. | 325. | 330. | 340. | 345. | 355. | 360. |
| 1960 | 1. | 8. | 47. | 100. | 204. | 270. | 291. | 293. | 321. | 318. | 320. | 344. | 349. | 370. | 379. | 375. | 380. |
| 1961 | 1. | 8. | 47. | 100. | 232. | 250. | 292. | 302. | 304. | 323. | 322. | 321. | 344. | 357. | 363. | 365. | 370. |
| 1962 | 1. | 8. | 47. | 100. | 219. | 291. | 300. | 316. | 324. | 326. | 335. | 338. | 334. | 347. | 354. | 358. | 358. |
| 1963 | 1. | 8. | 47. | 100. | 185. | 253. | 294. | 312. | 329. | 327. | 334. | 341. | 349. | 341. | 358. | 375. | 375. |
| 1964 | 1. | 8. | 47. | 100. | 194. | 213. | 264. | 317. | 363. | 353. | 349. | 354. | 357. | 359. | 365. | 402. | 402. |
| 1965 | 1. | 8. | 47. | 100. | 186. | 199. | 236. | 260. | 363. | 350. | 370. | 360. | 378. | 387. | 390. | 394. | 394. |
| 1966 | 1. | 8. | 47. | 100. | 185. | 219. | 222. | 249. | 306. | 354. | 377. | 391. | 379. | 378. | 361. | 383. | 383. |
| 1967 | 1. | 8. | 47. | 100. | 180. | 228. | 269. | 270. | 294. | 324. | 420. | 430. | 366. | 368. | 433. | 414. | 414. |
| 1968 | 1. | 8. | 47. | 100. | 115. | 206. | 266. | 275. | 274. | 285. | 350. | 325. | 363. | 408. | 388. | 378. | 378. |
| 1969 | 1. | 8. | 47. | 100. | 115. | 145. | 270. | 300. | 306. | 308. | 318. | 340. | 368. | 360. | 393. | 397. | 397. |
| 1970 | 1. | 8. | 47. | 100. | 209. | 272. | 230. | 295. | 317. | 323. | 325. | 329. | 380. | 370. | 380. | 391. | 391. |
| 1971 | 1. | 15. | 80. | 100. | 190. | 225. | 250. | 275. | 290. | 310. | 325. | 335. | 345. | 355. | 365. | 390. | 390. |
| 1972 | 1. | 10. | 70. | 150. | 150. | 140. | 210. | 240. | 270. | 300. | 325. | 335. | 345. | 355. | 365. | 390. | 390. |
| 1973 | 1. | 10. | 85. | 170. | 259. | 342. | 384. | 409. | 404. | 461. | 520. | 534. | 500. | 500. | 500. | 500. | 500. |
| 1974 | 1. | 10. | 85. | 170. | 259. | 342. | 384. | 409. | 444. | 461. | 520. | 543. | 482. | 482. | 482. | 482. | 482. |
| 1975 | 1. | 10. | 85. | 181. | 259. | 342. | 384. | 409. | 444. | 461. | 520. | 543. | 482. | 482. | 482. | 482. | 482. |
| 1976 | 1. | 10. | 85. | 181. | 259. | 342. | 384. | 409. | 444. | 461. | 520. | 543. | 482. | 482. | 482. | 482. | 482. |
| 1977 | 1. | 10. | 85. | 181. | 259. | 343. | 384. | 409. | 444. | 461. | 520. | 543. | 482. | 482. | 482. | 482. | 482. |
| 1978 | 1. | 10. | 85. | 180. | 294. | 326. | 371. | 409. | 461. | 476. | 520. | 543. | 500. | 500. | 500. | 500. | 500. |
| 1979 | 1. | 10. | 85. | 178. | 232. | 359. | 385. | 420. | 444. | 505. | 520. | 551. | 500. | 500. | 500. | 500. | 500. |
| 1980 | 1. | 10. | 85. | 175. | 283. | 347. | 402. | 421. | 465. | 465. | 520. | 534. | 500. | 500. | 500. | 500. | 500. |
| 1981 | 1. | 10. | 85. | 170. | 224. | 336. | 378. | 387. | 408. | 397. | 520. | 543. | 512. | 512. | 512. | 512. | 512. |
| 1982 | 1. | 10. | 85. | 170. | 204. | 303. | 355. | 383. | 395. | 413. | 453. | 468. | 506. | 506. | 506. | 506. | 506. |
| 1983 | 1. | 10. | 85. | 155. | 249. | 304. | 368. | 404. | 424. | 437. | 436. | 493. | 495. | 495. | 495. | 495. | 495. |
| 1984 | 1. | 10. | 85. | 140. | 204. | 295. | 338. | 376. | 395. | 407. | 413. | 422. | 437. | 437. | 437. | 437. | 437. |
| 1985 | 1. | 10. | 85. | 148. | 234. | 265. | 312. | 346. | 370. | 395. | 397. | 428. | 428. | 428. | 428. | 428. | 428. |
| 1986 | 1. | 10. | 85. | 54. | 206. | 265. | 289. | 339. | 368. | 391. | 382. | 388. | 395. | 395. | 395. | 395. | 395. |
| 1987 | 1. | 10. | 55. | 90. | 143. | 241. | 279. | 299. | 316. | 342. | 343. | 362. | 376. | 376. | 376. | 376. | 376. |
| 1988 | 1. | 15. | 50. | 98. | 135. | 197. | 277. | 315. | 339. | 343. | 359. | 365. | 376. | 376. | 376. | 376. | 376. |
| 1989 | 1. | 15. | 100. | 154. | 175. | 209. | 252. | 305. | 367. | 377. | 359. | 395. | 396. | 396. | 396. | 396. | 396. |
| 1990 | 1. | 8. | 48. | 219. | 198. | 258. | 288. | 309. | 428. | 370. | 403. | 387. | 440. | 440. | 440. | 440. | 440. |
| 1991 | 1. | 11. | 37. | 147. | 210. | 244. | 300. | 324. | 336. | 343. | 382. | 366. | 425. | 425. | 425. | 425. | 425. |
| 1992 | 1. | 7. | 30. | 128. | 224. | 296. | 327. | 355. | 345. | 367. | 341. | 361. | 430. | 470. | 470. | 470. | 450. |
| 1993 | 1. | 8. | 25. | 81. | 201. | 265. | 323. | 354. | 358. | 381. | 369. | 396. | 393. | 374. | 403. | 400. | 400. |
| 1994 | 1. | 10. | 25. | 75. | 151. | 254. | 318. | 371. | 347. | 412. | 382. | 407. | 410. | 410. | 410. | 410. | 410. |
| 1995 | 1. | 18. | 25. | 66. | 138. | 230. | 296. | 346. | 388. | 363. | 409. | 414. | 422. | 410. | 410. | 405. | 447. |
| 1996 | 1. | 18. | 25. | 76. | 118. | 188. | 261. | 316. | 346. | 374. | 390. | 390. | 384. | 398. | 398. | 398. | 398. |
| 1997 | 1. | 18. | 25. | 96. | 118. | 174. | 229. | 286. | 323. | 370. | 378. | 386. | 360. | 393. | 391. | 391. | 391. |
| 1998 | 1. | 18. | 25. | 74. | 147. | 174. | 217. | 242. | 278. | 304. | 310. | 359. | 340. | 344. | 385. | 363. | 375. |
| 1999 | 1. | 18. | 25. | 102. | 150. | 223. | 240. | 264. | 283. | 315. | 345. | 386. | 386. | 386. | 382. | 382. | 407. |
| 2000 | 1. | 18. | 25. | 102. | 150. | 223. | 240. | 264. | 283. | 315. | 345. | 386. | 386. | 386. | 382. | 382. | 407. |
| 2001 | 1. | 18. | 25. | 75. | 178. | 238. | 247. | 296. | 307. | 314. | 328. | 351. | 376. | 406. | 414. | 425. | 425. |
| 2002 | 1. | 10. | 23. | 57. | 177. | 241. | 275. | 302. | 311. | 314. | 328. | 341. | 372. | 405. | 415. | 467. | 409. |

# Table 3. Proportion of mature norwegian spring spawning herring used by SeaStar. 

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0.000 | 0.000 | 0.000 | 0.000 | 0.100 | 0.300 | 0.600 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1951 | 0.000 | 0.000 | 0.000 | 0.000 | 0.100 | 0.300 | 0.600 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 52 | 0.000 | 0.000 | 0.000 | 0.000 | 0.100 | 0.300 | 0.600 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1953 | 0.000 | 0.000 | 0.000 | 0.000 | 0.100 | 0.300 | 0.600 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | . 000 | 1.000 |
| 1954 | 0.000 | 0.000 | 0.000 | 0.000 | 0.100 | 0.300 | 0.600 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1955 | 0.000 | 0.000 | 0.000 | 0.080 | 0.220 | 0.370 | 0.850 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1956 | 0.000 | 0.000 | 0.000 | 0.080 | 0.220 | 0.370 | 0.850 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1957 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.500 | 0.600 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 58 | 0.000 | 0.000 | 0.000 | 0.080 | 0.220 | 0.370 | 0.850 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1959 | 0.000 | 0.000 | 0.000 | 0.080 | 0.220 | 0.370 | 0.850 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1960 | 0.000 | 0.000 | 0.000 | 0.080 | 0.220 | 0.370 | 0.850 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1961 | 0.000 | 0.000 | 0.000 | 0.040 | 0.350 | 0.680 | 0.940 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1962 | 0.000 | 0.000 | 0.000 | 0.000 | 0.110 | 0.670 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1963 | 0.000 | 0.000 | 0.000 | 0.040 | 0.030 | 0.320 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1964 | 0.000 | 0.000 | 0.000 | 0.020 | 0.060 | 0.280 | 0.320 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1965 | 0.000 | 0.000 | 0.000 | 0.000 | 0.340 | 0.350 | 0.760 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 66 | 0.000 | 0.000 | 0.000 | 0.010 | 0.150 | 1.000 | 0.960 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1967 | 0.000 | 0.000 | 0.000 | 0.000 | 0.010 | 0.230 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1968 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.010 | 0.760 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1969 | 0.000 | 0.000 | 0.000 | 0.620 | 0.890 | 0.950 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1970 | 0.000 | 0.000 | 0.000 | 0.060 | 0.130 | 0.310 | 0.170 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 71 | 0.000 | 0.000 | 0.000 | 0.100 | 0.250 | 0.600 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1972 | 0. | 0.000 | 0. | 0.000 | 0.100 | 0.250 | 0.600 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1973 | 0.000 | 0.000 | 0.000 | 0.500 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1974 | 0.000 | 0.000 | 0.000 | 0.500 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1975 | 0.000 | 0.000 | 0.000 | 0.500 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1976 | 0.000 | 0.000 | 0.000 | 0.500 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1977 | 0.0 | 0.000 | 0.000 | 0.730 | 0.890 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 19 | 0. | 0.000 | 0.000 | 0.130 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1979 | 0.000 | 0.000 | 0.000 | 0.100 | 0.620 | 0.950 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1980 | 0.000 | 0.000 | 0.000 | 0.250 | 0.500 | 0.970 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1981 | 0.000 | 0.000 | 0.000 | 0.300 | 0.500 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 82 | 0.000 | 0.000 | 0.000 | 0.100 | 0.480 | 0.700 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 83 | 0.000 | 0.000 | 0.000 | 0.100 | 0.500 | 0.690 | 0.710 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1984 | 0.000 | 0.000 | 0.000 | 0.100 | 0.500 | 0.900 | 0.950 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1985 | 0.000 | 0.000 | 0.000 | 0.100 | 0.500 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1986 | 0.000 | 0.000 | 0.000 | 0.100 | 0.200 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1987 | 0.000 | 0.000 | 0.000 | 0.100 | 0.300 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1988 | 0.000 | 0.000 | 0.000 | 0.100 | 0.300 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1989 | 0.000 | 0.000 | 0.000 | 0.100 | 0.300 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1990 | 0.000 | 0.000 | 0.000 | 0.400 | 0.800 | 0.900 | 0.900 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1991 | 0.000 | 0.000 | 0.000 | 0.100 | 0.700 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1992 | 0.000 | 0.000 | 0.000 | 0.100 | 0.200 | 0.800 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1993 | 0.000 | 0.000 | 0.000 | 0.010 | 0.300 | 0.800 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1994 | 0.000 | 0.000 | 0.000 | 0.010 | 0.300 | 0.800 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1995 | 0.000 | 0.000 | 0.000 | 0.000 | 0.300 | 0.800 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1996 | 0.000 | 0.000 | 0.000 | 0.000 | 0.300 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1997 | 0.000 | 0.000 | 0.000 | 0.000 | 0.300 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1998 | 0.000 | 0.000 | 0.000 | 0.000 | 0.300 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 1999 | 0.000 | 0.000 | 0.000 | 0.000 | 0.300 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.300 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.300 | 0.900 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 2002 | 0.0 | 0. | 0. | 0. | 0.300 | 0. | 1.000 | 1.000 | 1.000 | 1. | 1. | 1. | 1.000 | 1. | . |  | 1.000 |

Table 4. Stock in numbers by age from SeaStar (in billions)

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 693.066 | 23.975 | 15.621 | 8.289 | 2.985 | 3.711 | 6.585 | 5.947 | 8.468 | 3.213 | 2.645 | 2.882 | 5.417 | 4.743 | 0.559 | 1.161 | 2.574 |
| 1951 | 143.606 | 278.519 | 8.472 | 5.968 | 6.879 | 2.397 | 3.022 | 5.160 | 4.536 | 7.215 | 2.683 | 2.175 | 2.400 | 4.482 | 3.741 | 0.420 | 2.995 |
| 1952 | 96.420 | 57.343 | 108.387 | 3.190 | 3.619 | 5.564 | 1.903 | 2.449 | 3.963 | 3.345 | 6.138 | 2.233 | 1.776 | 1.966 | 3.622 | 2.897 | 2.656 |
| 1953 | 86.254 | 30.452 | 17.480 | 43.281 | 2.709 | 3.059 | 4.231 | 1.512 | 1.918 | 3.058 | 2.529 | 5.210 | 1.842 | 1.429 | 1.593 | 2.945 | 4.493 |
| 1954 | 44.367 | 31.435 | 9.158 | 6.736 | 36.565 | 2.288 | 2.539 | 3.311 | 1.225 | 1.548 | 2.341 | 1.810 | 4.427 | 1.501 | 1.142 | 1.279 | 5.924 |
| 1955 | 25.473 | 11.231 | 8.272 | 3.178 | 5.551 | 30.140 | 1.837 | 1.966 | 2.395 | 0.936 | 1.147 | 1.606 | 1.131 | 3.728 | 1.199 | 0.860 | 5.573 |
| 1956 | 30.646 | 7.057 | 2.735 | 3.038 | 2.649 | 4.521 | 24.045 | 1.475 | 1.517 | 1.807 | 0.726 | 0.808 | 1.108 | 0.785 | 3.154 | 0.953 | 5.094 |
| 1957 | 25.397 | 9.040 | 1.579 | 0.712 | 2.507 | 2.047 | 3.600 | 18.325 | 1.168 | 1.116 | 1.310 | 0.504 | 0.511 | 0.701 | 0.524 | 2.657 | 4.763 |
| 1958 | 23.095 | 7.136 | 1.577 | 0.502 | 0.591 | 1.811 | 1.619 | 2.886 | 13.931 | 0.938 | 0.843 | 0.958 | 0.352 | 0.328 | 0.465 | 0.329 | 6.156 |
| 1959 | 412.475 | 3.226 | 1.117 | 0.216 | 0.416 | 0.492 | 1.456 | 1.310 | 2.304 | 11.087 | 0.742 | 0.611 | 0.638 | 0.211 | 0.210 | 0.334 | 5.409 |
| 1960 | 197.510 | 156.288 | 1.185 | 0.247 | 0.172 | 0.333 | 0.400 | 1.117 | 1.021 | 1.760 | 8.519 | 0.556 | 0.411 | 0.366 | 0.100 | 0.109 | 4.803 |
| 1961 | 76.099 | 72.086 | 54.883 | 0.231 | 0.099 | 0.131 | 0.260 | 0.322 | 0.872 | 0.811 | 1.326 | 6.253 | 0.400 | 0.233 | 0.172 | 0.033 | 4.115 |
| 1962 | 19.006 | 26.982 | 19.058 | 20.474 | 0.170 | 0.078 | 0.109 | 0.210 | 0.259 | 0.694 | 0.652 | 1.015 | 4.707 | 0.298 | 0.159 | 0.090 | 3.535 |
| 1963 | 168.933 | 5.373 | 8.368 | 7.084 | 15.912 | 0.139 | 0.064 | 0.087 | 0.162 | 0.212 | 0.542 | 0.513 | 0.765 | 3.296 | 0.215 | 0.086 | 3.040 |
| 1964 | 93.902 | 65.618 | 0.833 | 2.098 | 5.392 | 12.920 | 0.115 | 0.054 | 0.072 | 0.123 | 0.174 | 0.367 | 0.356 | 0.497 | 1.980 | 0.112 | 2.574 |
| 1965 | 8.490 | 35.874 | 24.939 | 0.198 | 1.699 | 4.271 | 9.222 | 0.086 | 0.045 | 0.059 | 0.082 | 0.122 | 0.227 | 0.230 | 0.286 | 0.987 | 2.042 |
| 1966 | 51.405 | 1.983 | 12.174 | 8.320 | 0.087 | 1.225 | 3.146 | 5.897 | 0.056 | 0.025 | 0.044 | 0.053 | 0.068 | 0.102 | 0.098 | 0.117 | 2.441 |
| 1967 | 3.943 | 18.396 | 0.384 | 3.880 | 5.260 | 0.050 | 0.621 | 1.496 | 2.399 | 0.013 | 0.008 | 0.022 | 0.022 | 0.048 | 0.024 | 0.017 | 1.775 |
| 1968 | 5.184 | 1.331 | 1.181 | 0.111 | 2.048 | 1.509 | 0.018 | 0.144 | 0.238 | 0.469 | 0.003 | 0.002 | 0.015 | 0.011 | 0.033 | 0.004 | 1.459 |
| 1969 | 9.276 | 0.970 | 0.263 | 0.233 | 0.004 | 0.018 | 0.011 | 0.003 | 0.037 | 0.080 | 0.083 | 0.000 | 0.000 | 0.012 | 0.007 | 0.026 | 1.245 |
| 1970 | 0.661 | 3.414 | 0.071 | 0.016 | 0.026 | 0.002 | 0.007 | 0.005 | 0.002 | 0.021 | 0.038 | 0.038 | 0.000 | 0.000 | 0.011 | 0.006 | 1.093 |
| 1971 | 0.240 | 0.193 | 1.050 | 0.008 | 0.008 | 0.005 | 0.002 | 0.003 | 0.002 | 0.000 | 0.005 | 0.008 | 0.007 | 0.000 | 0.000 | 0.009 | 0.944 |
| 1972 | 2.376 | 0.078 | 0.051 | 0.373 | 0.005 | 0.006 | 0.003 | 0.000 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.820 |
| 1973 | 13.629 | 0.745 | 0.006 | 0.008 | 0.288 | 0.001 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.706 |
| 1974 | 8.631 | 5.523 | 0.301 | 0.001 | 0.004 | 0.225 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.607 |
| 1975 | 3.027 | 3.467 | 2.240 | 0.120 | 0.000 | 0.004 | 0.171 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.523 |
| 1976 | 8.080 | 1.211 | 1.407 | 0.910 | 0.100 | 0.000 | 0.002 | 0.118 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.450 |
| 1977 | 5.092 | 3.272 | 0.491 | 0.571 | 0.761 | 0.081 | 0.000 | 0.002 | 0.090 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.387 |
| 1978 | 6.200 | 2.043 | 1.326 | 0.198 | 0.471 | 0.633 | 0.069 | 0.000 | 0.001 | 0.067 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.333 |
| 1979 | 12.035 | 2.508 | 0.829 | 0.538 | 0.167 | 0.394 | 0.526 | 0.059 | 0.000 | 0.000 | 0.053 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.287 |
| 1980 | 1.473 | 4.872 | 1.017 | 0.336 | 0.458 | 0.142 | 0.333 | 0.443 | 0.050 | 0.000 | 0.000 | 0.043 | 0.000 | 0.000 | 0.000 | 0.000 | 0.247 |
| 1981 | 1.100 | 0.595 | 1.980 | 0.413 | 0.283 | 0.388 | 0.120 | 0.279 | 0.366 | 0.043 | 0.000 | 0.000 | 0.035 | 0.000 | 0.000 | 0.000 | 0.213 |
| 1982 | 2.329 | 0.442 | 0.241 | 0.798 | 0.352 | 0.239 | 0.326 | 0.102 | 0.236 | 0.308 | 0.037 | 0.000 | 0.000 | 0.029 | 0.000 | 0.000 | 0.183 |
| 1983 | 324.949 | 0.932 | 0.179 | 0.098 | 0.674 | 0.296 | 0.202 | 0.275 | 0.086 | 0.198 | 0.259 | 0.031 | 0.000 | 0.000 | 0.025 | 0.000 | 0.158 |
| 1984 | 11.528 | 132.034 | 0.376 | 0.072 | 0.081 | 0.560 | 0.246 | 0.168 | 0.230 | 0.072 | 0.167 | 0.216 | 0.027 | 0.000 | 0.000 | 0.021 | 0.136 |
| 1985 | 35.633 | 4.665 | 53.680 | 0.151 | 0.058 | 0.065 | 0.425 | 0.194 | 0.133 | 0.184 | 0.056 | 0.128 | 0.180 | 0.023 | 0.000 | 0.000 | 0.135 |
| 1986 | 6.041 | 14.469 | 1.888 | 21.692 | 0.110 | 0.035 | 0.041 | 0.245 | 0.113 | 0.063 | 0.100 | 0.039 | 0.082 | 0.109 | 0.020 | 0.000 | 0.113 |
| 1987 | 9.012 | 2.447 | 5.882 | 0.766 | 18.170 | 0.079 | 0.017 | 0.021 | 0.114 | 0.027 | 0.016 | 0.014 | 0.015 | 0.009 | 0.019 | 0.017 | 0.095 |
| 1988 | 27.873 | 3.655 | 0.991 | 2.369 | 0.641 | 15.174 | 0.050 | 0.011 | 0.011 | 0.072 | 0.012 | 0.005 | 0.008 | 0.006 | 0.002 | 0.010 | 0.097 |
| 1989 | 71.286 | 11.323 | 1.484 | 0.397 | 1.980 | 0.528 | 12.550 | 0.035 | 0.006 | 0.004 | 0.048 | 0.002 | 0.001 | 0.004 | 0.003 | 0.000 | 0.092 |
| 1990 | 127.096 | 28.978 | 4.602 | 0.587 | 0.339 | 1.701 | 0.449 | 10.501 | 0.027 | 0.005 | 0.003 | 0.039 | 0.000 | 0.000 | 0.003 | 0.002 | 0.079 |
| 1991 | 336.460 | 51.673 | 11.781 | 1.861 | 0.488 | 0.289 | 1.453 | 0.377 | 8.828 | 0.022 | 0.003 | 0.000 | 0.031 | 0.000 | 0.000 | 0.002 | 0.070 |
| 1992 | 378.721 | 136.794 | 21.006 | 4.788 | 1.594 | 0.418 | 0.248 | 1.237 | 0.316 | 7.395 | 0.016 | 0.002 | 0.000 | 0.026 | 0.000 | 0.000 | 0.062 |
| 1993 | 99.746 | 153.976 | 55.616 | 8.540 | 4.109 | 1.341 | 0.355 | 0.212 | 1.054 | 0.267 | 6.156 | 0.012 | 0.000 | 0.000 | 0.021 | 0.000 | 0.053 |
| 1994 | 32.632 | 40.549 | 62.602 | 22.607 | 7.324 | 3.438 | 1.074 | 0.297 | 0.179 | 0.879 | 0.212 | 4.918 | 0.010 | 0.000 | 0.000 | 0.018 | 0.046 |
| 1995 | 8.779 | 13.267 | 16.486 | 25.447 | 19.428 | 6.202 | 2.621 | 0.771 | 0.242 | 0.147 | 0.722 | 0.150 | 3.634 | 0.006 | 0.000 | 0.000 | 0.055 |
| 1996 | 72.998 | 3.569 | 5.394 | 6.702 | 21.849 | 16.401 | 4.760 | 1.664 | 0.449 | 0.194 | 0.112 | 0.557 | 0.051 | 2.282 | 0.001 | 0.000 | 0.047 |
| 1997 | 103.258 | 29.679 | 1.451 | 2.174 | 5.737 | 18.143 | 12.659 | 3.224 | 1.056 | 0.291 | 0.161 | 0.089 | 0.418 | 0.028 | 1.188 | 0.001 | 0.040 |
| 1998 | 201.620 | 41.981 | 12.067 | 0.576 | 1.750 | 4.686 | 13.950 | 9.046 | 2.069 | 0.606 | 0.194 | 0.120 | 0.047 | 0.276 | 0.006 | 0.679 | 0.036 |
| 1999 | 150.459 | 81.972 | 17.068 | 4.853 | 0.431 | 1.281 | 3.692 | 10.374 | 6.613 | 1.427 | 0.401 | 0.127 | 0.080 | 0.037 | 0.133 | 0.000 | 0.615 |
| 2000 | 17.643 | 61.172 | 33.327 | 6.936 | 4.049 | 0.337 | 0.978 | 2.779 | 7.440 | 4.612 | 0.958 | 0.247 | 0.096 | 0.032 | 0.025 | 0.032 | 0.470 |
| 2001 | 5.026 | 7.173 | 24.870 | 13.541 | 5.892 | 2.965 | 0.258 | 0.739 | 2.017 | 5.198 | 3.000 | 0.623 | 0.146 | 0.068 | 0.006 | 0.000 | 0.370 |
| 2002 | 159.329 | 2.043 | 2.916 | 10.110 | 11.560 | 4.922 | 2.156 | 0.186 | 0.547 | 1.461 | 3.696 | 2.112 | 0.468 | 0.103 | 0.055 | 0.002 | 0.298 |

# Table 5. Spawning stock biomass by age from SeaStar. 

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 1 | 12 | 13 | 14 | 15 | 16 | Sum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0.000 | 0.000 | 0.000 | 0.000 | 0.060 | 0.251 | 0.983 | 1.433 | 2.417 | 0.962 | 0.817 | 0.920 | 1.754 | 1.575 | 0.187 | 0.410 | 0.916 | 12.684 |
| 1951 | 0.000 | 0.000 | 0.000 | 0.000 | 0.137 | 0.162 | 0.453 | 1.244 | 1.276 | 2.165 | 0.830 | 0.693 | 0.776 | 1.492 | 1.258 | 0.148 | 1.063 | 11.696 |
| 1952 | 0.000 | 0.000 | 0.000 | 0.000 | 0.073 | 0.374 | 0.285 | 0.591 | 1.120 | 0.992 | 1.902 | 0.712 | 0.573 | 0.654 | 1.224 | 1.023 | 0.945 | 10.468 |
| 953 | 0.000 | 0.000 | 0.000 | 0.000 | 0.054 | 0.207 | 0.632 | 0.366 | 0.544 | 0.908 | 0.770 | 1.666 | 0.596 | 0.475 | 0.538 | 1.042 | 1.602 | 9.400 |
| 954 | 0.000 | 0.000 | 0.000 | 0.000 | 0.732 | 0.154 | 0.379 | 0.793 | 0.346 | 0.458 | 0.710 | 0.561 | 1.436 | 0.499 | 0.383 | 0.451 | 2.106 | 9.009 |
| 1955 | 0.000 | 0.000 | 0.000 | 0.025 | 0.233 | 2.322 | 0.397 | 0.527 | 0.675 | 0.278 | 0.349 | 0.503 | 0.360 | 1.247 | 0.404 | 0.302 | 1.976 | 9.599 |
| 1956 | 0.000 | 0.000 | 0.000 | 0.024 | 0.116 | 0.376 | 4.953 | 0.396 | 0.427 | 0.534 | 0.221 | 0.251 | 0.349 | 0.256 | 1.070 | 0.336 | 1.812 | 11.121 |
| 1957 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.228 | 0.539 | 4.671 | 0.331 | 0.331 | 0.400 | 0.158 | 0.161 | 0.229 | 0.173 | 0.934 | 1.688 | 9.842 |
| 958 | 0.000 | 0.000 | 0.000 | 0.004 | 0.026 | 0.159 | 0.393 | 0.833 | 3.990 | 0.279 | 0.257 | 0.304 | 0.114 | 0.108 | 0.158 | 0.114 | 2.157 | 8.895 |
| 959 | 0.000 | 0.000 | 0.000 | 0.002 | 0.018 | 0.045 | 0.313 | 0.371 | 0.673 | 3.294 | 0.227 | 0.191 | 0.199 | 0.067 | 0.068 | 0.113 | 1.858 | 7.438 |
| 1960 | 0.000 | 0.000 | 0.000 | 0.002 | 0.008 | 0.032 | 0.097 | 0.319 | 0.320 | 0.544 | 2.643 | 0.185 | 0.135 | 0.126 | 0.034 | 0.038 | 1.690 | 6.173 |
| 1961 | 0.000 | 0.000 | 0.000 | 0.000 | 0.008 | 0.022 | 0.070 | 0.095 | 0.259 | 0.256 | 0.416 | 1.951 | 0.134 | 0.080 | 0.059 | 0.011 | 1.332 | 4.693 |
| 1962 | 0.000 | 0.000 | 0.000 | 0.000 | 0.004 | 0.015 | 0.032 | 0.065 | 0.082 | 0.221 | 0.213 | 0.333 | 1.517 | 0.100 | 0.053 | 0.027 | 1.075 | 3.738 |
| 1963 | 0.000 | 0.000 | 0.000 | 0.028 | 0.086 | 0.011 | 0.017 | 0.027 | 0.052 | 0.068 | 0.174 | 0.169 | 0.256 | 1.068 | 0.072 | 0.028 | 0.980 | 3.035 |
| 1964 | 0.000 | 0.000 | 0.000 | 0.004 | 0.061 | 0.745 | 0.009 | 0.017 | 0.026 | 0.042 | 0.058 | 0.124 | 0.122 | 0.169 | 0.674 | 0.042 | 0.962 | 3.054 |
| 1965 | 0.000 | 0.000 | 0.000 | 0.000 | 0.104 | 0.289 | 1.582 | 0.021 | 0.015 | 0.020 | 0.029 | 0.041 | 0.079 | 0.082 | 0.102 | 0.333 | 0.689 | 3.386 |
| 1966 | 0.000 | 0.000 | 0.000 | 0.008 | 0.002 | 0.251 | 0.622 | 1.342 | 0.015 | 0.008 | 0.015 | 0.019 | 0.025 | 0.033 | 0.030 | 0.036 | 0.752 | 3.159 |
| 1967 | 0.000 | 0.000 | 0.000 | 0.000 | 0.008 | 0.002 | 0.144 | 0.336 | 0.599 | 0.004 | 0.003 | 0.009 | 0.007 | 0.017 | 0.009 | 0.006 | 0.586 | 1.731 |
| 1968 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 | 0.003 | 0.035 | 0.058 | 0.112 | 0.000 | 0.000 | 0.005 | 0.004 | 0.013 | 0.001 | 0.51 | 0.747 |
| 1969 | 0.000 | 0.000 | 0.000 | 0.012 | 0.000 | 0.002 | 0.003 | 0.000 | 0.011 | 0.023 | 0.025 | 0.000 | 0.000 | 0.004 | 0.003 | 0.010 | 0.486 | 0.580 |
| 1970 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.006 | 0.011 | 0.011 | 0.000 | 0.000 | 0.004 | 0.002 | 0.420 | 0.456 |
| 1971 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.002 | 0.002 | 0.000 | 0.000 | 0.003 | 0.362 | 0.374 |
| 1972 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.311 | 0.313 |
| 973 | 0.000 | 0.000 | 0.000 | 0.000 | . 065 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.34 | 0.412 |
| 974 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.075 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.28 | 0.363 |
| 1975 | 0.000 | 0.000 | 0.000 | 0.011 | 0.000 | 0.001 | 0.063 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.247 | 0.322 |
| 1976 | 0.000 | 0.000 | 0.000 | 0.081 | 0.023 | 0.000 | 0.000 | 0.047 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.213 | 0.365 |
| 1977 | 0.000 | 0.000 | 0.000 | 0.074 | 0.172 | 0.027 | 0.000 | 0.000 | 0.039 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.184 | 0.497 |
| 78 | 0.000 | 0.000 | 0.000 | 0.005 | 0.123 | 0.203 | 0.025 | 0.000 | 0.000 | 0.031 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.164 | 0.551 |
| 1979 | 0.000 | 0.000 | 0.00 | 0.009 | 0.024 | 0. | 0.199 | 0.024 | 0.000 | 0.0 | 0.027 | 0. | 0.000 | 0.000 | 0.000 | 0.000 | 0. | 0.557 |
| 1980 | 0.000 | 0.000 | 0.000 | 0.014 | 0.064 | 0.047 | 0.132 | 0.183 | 0.023 | 0.000 | 0.000 | 0.023 | 0.000 | 0.000 | 0.000 | 0.000 | 0.11 | 0.603 |
| 1981 | 0.000 | 0.000 | 0.000 | 0.021 | 0.031 | 0.115 | 0.045 | 0.106 | 0.147 | 0.017 | 0.000 | 0.000 | 0.018 | 0.000 | 0.000 | 0.000 | 0.093 | 0.593 |
| 1982 | 0.000 | 0.000 | 0.000 | 0.013 | 0.034 | 0.050 | 0.114 | 0.038 | 0.092 | 0.125 | 0.016 | 0.000 | 0.000 | 0.015 | 0.000 | 0.000 | 0.088 | 0.584 |
| 1983 | 0.000 | 0.000 | 0.000 | 0.001 | 0.082 | 0.061 | 0.052 | 0.109 | 0.036 | 0.085 | 0.111 | 0.015 | 0.000 | 0.000 | 0.012 | 0.000 | 0.074 | 0.639 |
| 984 | 0.000 | 0.0 | 0.000 | 0.000 | 0.008 | 0.145 | 0.077 | 0.062 | 0.089 | 0.029 | 0.067 | 0.090 | 0.012 | 0.000 | 0.000 | 0.009 | 0.058 | 0.645 |
| 5 | 0.000 | 0.000 | 0.000 | 0.002 | 0.006 | 0.015 | 0.126 | 0.064 | 0.046 | 0.068 | 0.021 | 0.052 | 0.073 | 0.010 | 0.000 | 0.000 | 0.056 | 0.539 |
| 1986 | 0.000 | 0.000 | 0.000 | 0.115 | 0.004 | 0.008 | 0.011 | 0.077 | 0.036 | 0.022 | 0.031 | 0.014 | 0.026 | 0.036 | 0.008 | 0.000 | 0.04 | 0.432 |
| 1987 | 0.000 | 0.000 | 0.000 | 0.007 | 0.766 | 0.016 | 0.005 | 0.006 | 0.034 | 0.009 | 0.005 | 0.005 | 0.005 | 0.003 | 0.007 | 0.006 | 0.035 | 0.908 |
| 1988 | 0.000 | 0.000 | 0.000 | 0.023 | 0.025 | 2.640 | 0.013 | 0.003 | 0.003 | 0.024 | 0.004 | 0.001 | 0.003 | 0.002 | 0.000 | 0.004 | 0.036 | 2.782 |
| 1989 | 0.000 | 0.000 | 0.000 | 0.006 | 0.102 | 0.098 | 3.107 | 0.010 | 0.002 | 0.001 | 0.017 | 0.000 | 0.000 | 0.002 | 0.000 | 0.000 | 0.036 | 3.383 |
| 1990 | 0.000 | 0.000 | 0.000 | 0.051 | 0.053 | 0.389 | 0.114 | 2.870 | 0.011 | 0.002 | 0.000 | 0.015 | 0.000 | 0.000 | 0.001 | 0.000 | 0.034 | 3.542 |
| 91 | 0.000 | 0.000 | 0.000 | 0.027 | 0.071 | 0.070 | 0.429 | 0.120 | 2.914 | 0.007 | 0.000 | 0.000 | 0.013 | 0.000 | 0.000 | 0.000 | 0.028 | 3.681 |
| 1992 | 0.000 | 0.000 | 0.000 | 0.060 | 0.070 | 0.097 | 0.080 | 0.432 | 0.107 | 2.665 | 0.005 | 0.000 | 0.000 | 0.012 | 0.000 | 0.000 | 0.027 | 3.557 |
| 1993 | 0.000 | 0.000 | 0.000 | 0.007 | 0.243 | 0.278 | 0.113 | 0.074 | 0.370 | 0.099 | 2.221 | 0.005 | 0.000 | 0.000 | 0.008 | 0.000 | 0.021 | 3.440 |
| 1994 | 0.000 | 0.000 | 0.000 | 0.017 | 0.326 | 0.680 | 0.330 | 0.108 | 0.061 | 0.355 | 0.078 | 1.942 | 0.004 | 0.000 | 0.000 | 0.007 | 0.018 | 3.928 |
| 1995 | 0.000 | 0.000 | 0.000 | 0.000 | 0.791 | 1.111 | 0.741 | 0.253 | 0.092 | 0.052 | 0.288 | 0.056 | 1.464 | 0.002 | 0.000 | 0.000 | 0.024 | 4.873 |
| 1996 | 0.000 | 0.000 | 0.000 | 0.000 | 0.759 | 2.704 | 1.195 | 0.503 | 0.149 | 0.071 | 0.043 | 0.211 | 0.018 | 0.851 | 0.000 | 0.000 | 0.018 | 6.522 |
| 1997 | 0.000 | 0.000 | 0.000 | 0.000 | 0.199 | 2.768 | 2.803 | 0.882 | 0.323 | 0.103 | 0.059 | 0.032 | 0.144 | 0.009 | 0.439 | 0.000 | 0.015 | 7.778 |
| 1998 | 0.000 | 0.000 | 0.000 | 0.000 | 0.075 | 0.717 | 2.939 | 2.122 | 0.554 | 0.177 | 0.058 | 0.041 | 0.016 | 0.088 | 0.001 | 0.238 | 0.013 | 7.038 |
| 1999 | 0.000 | 0.000 | 0.000 | 0.000 | 0.019 | 0.250 | 0.861 | 2.649 | 1.805 | 0.432 | 0.132 | 0.048 | 0.028 | 0.014 | 0.044 | 0.000 | 0.242 | 6.525 |
| 2000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.177 | 0.066 | 0.228 | 0.711 | 2.031 | 1.392 | 0.317 | 0.090 | 0.036 | 0.010 | 0.004 | 0.012 | 0.185 | 5.259 |
| 2001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.309 | 0.615 | 0.062 | 0.212 | 0.599 | 1.577 | 0.950 | 0.213 | 0.053 | 0.027 | 0.002 | 0.000 | 0.153 | 4.773 |
| 2002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.601 | 1.046 | 0.574 | 0.054 | 0.164 | 0.442 | 1.169 | 0.696 | 0.169 | 0.041 | 0.022 | 0.001 | 0.118 | 5.098 |

# Table 6. Fishing mortality for norwegian spring spawning herring from SeaStar 

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0.012 | 0.140 | 0.062 | 0.037 | 0.069 | 0.055 | 0.094 | 0.121 | 0.010 | 0.030 | 0.046 | 0.033 | 0.039 | 0.087 | 0.137 | 0.104 | 0.104 |
| 1951 | 0.018 | 0.044 | 0.077 | 0.350 | 0.062 | 0.081 | 0.060 | 0.114 | 0.154 | 0.012 | 0.034 | 0.052 | 0.050 | 0.063 | 0.106 | 0.130 | 0.130 |
| 1952 | 0.253 | 0.288 | 0.018 | 0.013 | 0.018 | 0.124 | 0.080 | 0.094 | 0.109 | 0.130 | 0.014 | 0.042 | 0.068 | 0.060 | 0.05 | 0.100 | 0.100 |
| 1953 | 0.109 | 0.302 | 0.054 | 0.019 | 0.019 | 0.036 | 0.095 | 0.060 | 0.064 | 0.117 | 0.184 | 0.013 | 0.055 | 0.074 | 0.069 | 0.082 | 0.082 |
| 1954 | 0.474 | 0.435 | 0.158 | 0.044 | 0.043 | 0.070 | 0.106 | 0.174 | 0.120 | 0.150 | 0.227 | 0.321 | 0.022 | 0.075 | 0.134 | 0.113 | 0.113 |
| 1955 | 0.384 | 0.512 | 0.102 | 0.032 | 0.055 | 0.076 | 0.069 | 0.110 | 0.132 | 0.103 | 0.201 | 0.221 | 0.215 | 0.017 | 0.079 | 0.139 | 0.139 |
| 56 | 0.321 | 0.597 | 0.446 | 0.042 | 0.108 | 0.078 | 0.122 | 0.084 | 0.157 | 0.172 | 0.216 | 0.307 | 0.308 | 0.254 | 0.022 | 0.106 | 0.106 |
| 1957 | 0.369 | 0.846 | 0.246 | 0.036 | 0.175 | 0.084 | 0.071 | 0.124 | 0.069 | 0.131 | 0.163 | 0.210 | 0.295 | 0.261 | 0.316 | 0.14 | 0.147 |
| 1958 | 1.068 | 0.954 | 1.087 | 0.038 | 0.033 | 0.068 | 0.061 | 0.075 | 0.078 | 0.085 | 0.171 | 0.256 | 0.360 | 0.294 | 0.180 | 0.25 | 0.257 |
| 1959 | 0.070 | 0.102 | 0.610 | 0.078 | 0.072 | 0.058 | 0.115 | 0.099 | 0.119 | 0.114 | 0.138 | 0.248 | 0.407 | 0.601 | 0.506 | 0.320 | 0.320 |
| 1960 | 0.108 | 0.147 | 0.733 | 0.759 | 0.121 | 0.095 | 0.068 | 0.097 | 0.081 | 0.133 | 0.159 | 0.180 | 0.416 | 0.602 | 0.951 | 0.620 | 0.620 |
| 1961 | 0.137 | 0.430 | 0.086 | 0.157 | 0.092 | 0.034 | 0.064 | 0.067 | 0.079 | 0.068 | 0.117 | 0.134 | 0.144 | 0.233 | 0.501 | 1.188 | 1.188 |
| 1962 | 0.363 | 0.271 | 0.090 | 0.102 | 0.052 | 0.044 | 0.074 | 0.109 | 0.051 | 0.096 | 0.091 | 0.133 | 0.206 | 0.174 | 0.464 | 1.478 | 1.478 |
| 1963 | 0.046 | 0.964 | 0.483 | 0.123 | 0.058 | 0.042 | 0.031 | 0.045 | 0.130 | 0.049 | 0.241 | 0.216 | 0.282 | 0.360 | 0.508 | 1.362 | 1.362 |
| 1964 | 0.062 | 0.067 | 0.536 | 0.061 | 0.083 | 0.187 | 0.138 | 0.031 | 0.046 | 0.247 | 0.201 | 0.330 | 0.287 | 0.403 | 0.546 | 0.578 | 0.578 |
| 65 | 0.554 | 0.181 | 0.198 | 0.671 | 0.177 | 0.156 | 0.297 | 0.280 | 0.444 | 0.145 | 0.287 | 0.435 | 0.649 | 0.705 | 0.741 | 1.406 | 1.406 |
| 1966 | 0.128 | 0.743 | 0.244 | 0.308 | 0.404 | 0.529 | 0.593 | 0.749 | 1.305 | 0.974 | 0.557 | 0.755 | 0.191 | 1.307 | 1.591 | 2.028 | 2.028 |
| 1967 | 0.186 | 1.845 | 0.339 | 0.489 | 1.099 | 0.849 | 1.312 | 1.690 | 1.483 | 1.322 | 1.444 | 0.192 | 0.553 | 0.221 | 1.573 | 2.106 | 2.106 |
| 1968 | 0.776 | 0.723 | 0.725 | 3.253 | 4.593 | 4.741 | 1.770 | 1.216 | 0.937 | 1.577 | 1.267 | 1.294 | 0.058 | 0.291 | 0.088 | 0.603 | 0.603 |
| 1969 | 0.100 | 1.713 | 1.882 | 2.053 | 0.265 | 0.759 | 0.592 | 0.327 | 0.420 | 0.601 | 0.627 | 0.587 | 0.820 | 0.017 | 0.032 | 0.017 | 0.017 |
| 1970 | 0.332 | 0.279 | 1.317 | 0.541 | 1.513 | 0.308 | 0.683 | 1.076 | 1.027 | 1.187 | 1.375 | 1.562 | 2.640 | 1.338 | 0.021 | 0.01 | 0.019 |
| 1971 | 0.222 | 0.429 | 0.136 | 0.292 | 0.145 | 0.321 | 0.289 | 0.483 | 1.466 | 1.384 | 2.044 | 2.355 | 1.890 | 0.052 | 1.060 | 0.015 | 0.015 |
| 1972 | 0.260 | 1.730 | 0.985 | 0.108 | 1.393 | 1.012 | 2.107 | 1.392 | 3.372 | 1.134 | 0.010 | 2.075 | 3.025 | 1.239 | 0.064 | 0.116 | 0.116 |
| 1973 | 0.003 | 0.007 | 0.642 | 0.404 | 0.099 | 1.075 | 1.930 | 2.817 | 2.316 | 0.022 | 0.012 | 0.011 | 0.017 | 0.039 | 2.003 | 0.079 | 0.079 |
| 1974 | 0.012 | 0.002 | 0.021 | 0.094 | 0.060 | 0.125 | 2.162 | 2.209 | 0.068 | 0.061 | 0.026 | 0.015 | 0.013 | 0.020 | 0.048 | 0.042 | 0.042 |
| 75 | 0.016 | 0.002 | 0.001 | 0.030 | 0.163 | 0.317 | 0.215 | 0.191 | 0.101 | 0.085 | 0.077 | 0.032 | 0.017 | 0.016 | 0.023 | 0.058 | 0.058 |
| 1976 | 0.004 | 0.003 | 0.001 | 0.028 | 0.060 | 0.002 | 0.000 | 0.127 | 0.050 | 0.064 | 0.109 | 0.097 | 0.038 | 0.020 | 0.019 | 0.028 | 0.028 |
| 1977 | 0.013 | 0.003 | 0.010 | 0.043 | 0.034 | 0.004 | 0.002 | 0.264 | 0.139 | 0.062 | 0.080 | 0.144 | 0.126 | 0.046 | 0.024 | 0.022 | 0.022 |
| 1978 | 0.005 | 0.002 | 0.001 | 0.017 | 0.028 | 0.035 | 0.014 | 0.002 | 0.732 | 0.084 | 0.077 | 0.101 | 0.198 | 0.169 | 0.057 | 0.029 | 0.029 |
| 1979 | 0.004 | 0.002 | 0.004 | 0.013 | 0.012 | 0.019 | 0.023 | 0.006 | 0.002 | 0.002 | 0.053 | 0.097 | 0.132 | 0.294 | 0.242 | 0.070 | 0.070 |
| 80 | 0.007 | 0.000 | 0.000 | 0.021 | 0.014 | 0.017 | 0.027 | 0.039 | 0.009 | 0.023 | 0.002 | 0.069 | 0.126 | 0.179 | 0.506 | 0.379 | 0.379 |
| 81 | 0.012 | 0.003 | 0.009 | 0.011 | 0.018 | 0.024 | 0.020 | 0.018 | 0.025 | 0.009 | 0.433 | 0.375 | 0.030 | 0.170 | 0.259 | 1.396 | 1.396 |
| 1982 | 0.015 | 0.004 | 0.001 | 0.019 | 0.024 | 0.020 | 0.021 | 0.021 | 0.023 | 0.021 | 0.004 | 0.256 | 0.188 | 0.004 | 0.242 | 0.417 | 0.417 |
| 1983 | 0.000 | 0.008 | 0.015 | 0.036 | 0.034 | 0.035 | 0.034 | 0.027 | 0.016 | 0.025 | 0.031 | 0.005 | 0.457 | 2.668 | 0.038 | 0.381 | 0.381 |
| 1984 | 0.005 | 0.000 | 0.010 | 0.070 | 0.074 | 0.126 | 0.083 | 0.085 | 0.076 | 0.114 | 0.112 | 0.033 | 0.000 | 0.017 | 0.115 | 0.090 | 0.090 |
| 19 | 0.001 | 0.004 | 0.006 | 0.166 | 0.343 | 0.320 | 0.400 | 0.396 | 0.591 | 0.461 | 0.215 | 0.302 | 0.356 | 0.000 | 0.020 | 0.153 | 0.153 |
| 1986 | 0.00 | 0.0 | 0. | 0.027 | 0.189 | 0.588 | 0.530 | 0.619 | 1.265 | 1.253 | 1.781 | 0.783 | 2.057 | 1.578 | 0.000 | 0.024 | 0.024 |
| 7 | 0.00 | 0.004 | 0.010 | 0.028 | 0.030 | 0.296 | 0.254 | 0.462 | 0.309 | 0.640 | 1.071 | 0.408 | 0.811 | 1.517 | 0.496 | 0.029 | 0.029 |
| 1988 | 0.000 | 0.001 | 0.015 | 0.029 | 0.043 | 0.040 | 0.226 | 0.436 | 0.859 | 0.247 | 1.468 | 1.081 | 0.574 | 0.688 | 4.860 | 0.059 | 0.059 |
| 1989 | 0.000 | 0.000 | 0.027 | 0.008 | 0.002 | 0.012 | 0.028 | 0.115 | 0.149 | 0.198 | 0.076 | 0.921 | 0.789 | 0.090 | 0.118 | 0.100 | 0.100 |
| 1990 | 0.000 | 0.000 | 0.005 | 0.035 | 0.008 | 0.008 | 0.026 | 0.024 | 0.054 | 0.436 | 1.447 | 0.070 | 1.744 | 0.458 | 0.225 | 0.100 | 0.100 |
| 1991 | 0.000 | 0.000 | 0.000 | 0.005 | 0.006 | 0.005 | 0.011 | 0.026 | 0.027 | 0.133 | 0.214 | 0.176 | 0.024 | 2.075 | 3.965 | 0.303 | 0.303 |
| 1992 | 0.000 | 0.000 | 0.000 | 0.003 | 0.023 | 0.013 | 0.005 | 0.010 | 0.020 | 0.033 | 0.179 | 0.486 | 1.007 | 0.053 | 0.082 | 0.262 | 0.262 |
| 1993 | 0.000 | 0.000 | 0.000 | 0.004 | 0.028 | 0.073 | 0.027 | 0.019 | 0.031 | 0.078 | 0.075 | 0.000 | 0.001 | 0.008 | 0.000 | 0.104 | 0.104 |
| 4 | 0.000 | 0.000 | 0.000 | 0.002 | 0.016 | 0.121 | 0.181 | 0.058 | 0.050 | 0.047 | 0.200 | 0.153 | 0.360 | 0.938 | 3.156 | 0.130 | 0.130 |
| 1995 | 0.000 | 0.000 | 0.000 | 0.002 | 0.019 | 0.115 | 0.304 | 0.390 | 0.072 | 0.124 | 0.110 | 0.925 | 0.315 | 1.287 | 4.046 | 0.301 | 0.301 |
| 1996 | 0.000 | 0.000 | 0.009 | 0.006 | 0.036 | 0.109 | 0.240 | 0.305 | 0.285 | 0.032 | 0.074 | 0.137 | 0.463 | 0.503 | 0.000 | 0.301 | 0.301 |
| 1997 | 0.000 | 0.000 | 0.024 | 0.067 | 0.052 | 0.113 | 0.186 | 0.294 | 0.406 | 0.256 | 0.144 | 0.497 | 0.266 | 1.366 | 0.409 | 0.301 | 0.301 |
| 1998 | 0.000 | 0.000 | 0.011 | 0.141 | 0.162 | 0.089 | 0.146 | 0.163 | 0.222 | 0.263 | 0.270 | 0.258 | 0.084 | 0.580 | 6.687 | 0.188 | 0.188 |
| 1999 | 0.000 | 0.000 | 0.000 | 0.031 | 0.094 | 0.120 | 0.134 | 0.182 | 0.210 | 0.249 | 0.336 | 0.131 | 0.775 | 0.236 | 1.267 | 0.180 | 0.180 |
| 2000 | 0.000 | 0.000 | 0.000 | 0.013 | 0.162 | 0.118 | 0.130 | 0.171 | 0.209 | 0.280 | 0.280 | 0.375 | 0.201 | 1.476 | 8.015 | 0.171 | 0.171 |
| 2001 | 0.000 | 0.000 | 0.000 | 0.008 | 0.030 | 0.169 | 0.177 | 0.151 | 0.173 | 0.191 | 0.201 | 0.136 | 0.193 | 0.057 | 0.867 | 0.163 | 0.163 |
| 002 | 0.000 | 0. | 0. | 0.021 | 0.061 | 0. | 0.175 | 0.189 | 0.202 | 0.215 | 0.212 | 0.188 | 0.129 | 0. | 0.146 | 0.154 | 0. |

Table 7. Summary table SeaStar. (Biomasses in million tonnes, Recruits in billions)

|  | Recruits | Total | Spawning stock | Fbar |
| :---: | :---: | :---: | :---: | :---: |
|  | Age 0 | biomass | biomass | 5-14 |
| 1950 | 693.066 | 17.383 | 12.684 | 0.060 |
| 1951 | 143.606 | 17.430 | 11.696 | 0.073 |
| 1952 | 96.420 | 18.496 | 10.468 | 0.078 |
| 1953 | 86.254 | 16.558 | 9.400 | 0.073 |
| 1954 | 44.367 | 18.087 | 9.009 | 0.129 |
| 1955 | 25.473 | 15.605 | 9.599 | 0.088 |
| 1956 | 30.646 | 13.902 | 11.121 | 0.124 |
| 1957 | 25.397 | 11.313 | 9.842 | 0.126 |
| 1958 | 23.095 | 9.799 | 8.895 | 0.096 |
| 1959 | 412.475 | 8.406 | 7.438 | 0.137 |
| 1960 | 197.510 | 8.084 | 6.173 | 0.166 |
| 1961 | 76.099 | 8.264 | 4.693 | 0.126 |
| 1962 | 19.006 | 7.238 | 3.738 | 0.172 |
| 1963 | 168.933 | 7.457 | 3.035 | 0.299 |
| 1964 | 93.902 | 7.081 | 3.054 | 0.241 |
| 1965 | 8.490 | 6.443 | 3.386 | 0.277 |
| 1966 | 51.405 | 5.065 | 3.159 | 0.690 |
| 1967 | 3.943 | 3.587 | 1.731 | 1.496 |
| 1968 | 5.184 | 1.450 | 0.747 | 3.414 |
| 1969 | 9.276 | 0.634 | 0.580 | 0.547 |
| 1970 | 0.661 | 0.508 | 0.456 | 1.211 |
| 1971 | 0.240 | 0.471 | 0.374 | 1.554 |
| 1972 | 2.376 | 0.386 | 0.313 | 1.689 |
| 1973 | 13.629 | 0.452 | 0.412 | 1.645 |
| 1974 | 8.631 | 0.461 | 0.363 | 0.130 |
| 1975 | 3.027 | 0.569 | 0.322 | 0.217 |
| 1976 | 8.080 | 0.597 | 0.365 | 0.124 |
| 1977 | 5.092 | 0.636 | 0.497 | 0.077 |
| 1978 | 6.200 | 0.745 | 0.551 | 0.039 |
| 1979 | 12.035 | 0.783 | 0.557 | 0.022 |
| 1980 | 1.473 | 0.865 | 0.603 | 0.032 |
| 1981 | 1.100 | 0.887 | 0.593 | 0.022 |
| 1982 | 2.329 | 0.806 | 0.584 | 0.021 |
| 1983 | 324.949 | 1.150 | 0.639 | 0.030 |
| 1984 | 11.528 | 2.063 | 0.645 | 0.093 |
| 1985 | 35.633 | 5.241 | 0.539 | 0.393 |
| 1986 | 6.041 | 1.857 | 0.432 | 1.139 |
| 1987 | 9.012 | 3.169 | 0.908 | 0.447 |
| 1988 | 27.873 | 3.538 | 2.782 | 0.045 |
| 1989 | 71.286 | 4.142 | 3.383 | 0.028 |
| 1990 | 127.096 | 4.656 | 3.542 | 0.022 |
| 1991 | 336.460 | 5.364 | 3.681 | 0.025 |
| 1992 | 378.721 | 6.450 | 3.557 | 0.029 |
| 1993 | 99.746 | 7.570 | 3.440 | 0.066 |
| 1994 | 32.632 | 8.668 | 3.928 | 0.134 |
| 1995 | 8.779 | 9.555 | 4.873 | 0.222 |
| 1996 | 72.998 | 9.647 | 6.522 | 0.182 |
| 1997 | 103.258 | 9.723 | 7.778 | 0.176 |
| 1998 | 201.620 | 8.826 | 7.038 | 0.156 |
| 1999 | 150.459 | 9.378 | 6.525 | 0.194 |
| 2000 | 17.643 | 8.557 | 5.259 | 0.235 |
| 2001 | 5.026 | 7.505 | 4.773 | 0.181 |
| 2002 | 159.329 | 7.628 | 5.098 | 0.152 |

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### 1.1 Terms of reference

The Northern Pelagic and Blue Whiting Fisheries Working Group [WGNPBW] (Chair: A. Gudmundsdottir, Iceland) will meet in ICES Headquarters from 29 April to 8 May 2003 to:
a) assess the status of and provide catch options for 2004 for the Norwegian spring-spawning herring stock;
b) assess the status of and provide catch options for 2004 for the blue whiting stock;
c) assess the status of and provide catch options for the 2003-2004 season for the Icelandic summer-spawning herring stocks;
d) assess the status of capelin in Sub-areas V and XIV and provide catch options for the summer/autumn 2003 and winter 2004 seasons;
e) assess the status of and provide catch options for capelin in Sub-areas I and II (excluding Division IIa west of $5^{\circ} \mathrm{W}$ ) in 2004;
f) provide as detailed information as possible on the age/size composition in different segments of the blue whiting fishery;
g) provide information on the species compositions in those fisheries that take appreciable amounts of blue whiting, and on the age/size composition by species of these catches [EC request for information on the industrial fisheries];
h) evaluate the effect on the blue whiting stock and the fisheries of possible measures to reduce exploitation of juveniles. The evaluation should include but not be restricted to the effects of introducing a minimum size and closed areas/seasons;
i) continue the evaluation of candidates of harvest control rules for blue whiting;
j) provide specific information on possible deficiencies in the assessments including at least: Major inadequacies in the data on catches, effort or discards; major inadequacies if any in research vessel surveys data and major difficulties if any in model formulation; including inadequacies in available software. The Group should clarify the consequences from these deficiencies for $a$ ) assessment of the status of the stocks and $b$ ) for the projection;
k) for stocks for which a full analytical assessment is presented, comment on this meeting's assessments compared to the last assessment of the same stock;

1) comment on the PA reference points proposed by the Study Group on Precautionary Reference Points for Advice on Fishery Management;
m) structure the assessment report following the guidelines as adopted by ACFM in October 2002 with special attention to the quality issues.

WGNPBW will report by 9 May 2003 for the attention of ACFM.

Government of Russia on behalf of the coastal states (Iceland,
Norway, Faroe
Islands, Russia and
EC)
22.10.2002

At the Fisheries Consultations on the management of Norwegian spring-spawning (Atlanto-Scandian) herring stock in the North-East Atlantic for 2003 the Parties agreed to submit the following request to ICES:

1. ICES is requested to evaluate and compare the models ISVPA and SeaStar and to comment on their applicability to assess the state of the Norwegian springspawning (Atlanto-Scandian) herring stock.
2. Based on the evaluation of the models and any new information and taking into account the long-term management plan agreed by the Parties, ICES is requested to review its TAC advice for 2003.
ICES is requested to report to the Parties as soon as possible, and not later than the first half of June 2003.

## EC, The Faroe Islands, Greenland, Iceland, Norway and Russia

21.11.2002

As an outcome of the coastal state meeting on blue whiting 7-8 November:

1. The parties noted that similar discrepancies exist for the assessment of the Norwegian Spring-spawning (Atlanto-Scandian) herring stock and that a request to ICES to evaluate the two assessment models with respect to Norwegian Spring-spawning herring has been put forward by the Russian Federation on behalf of the coastal states. The Parties request ICES to extend these evaluations to also include assessment of blue whiting.
2. The Parties further noted that there is a lack of co-ordination on research on blue whiting. ICES is therefore invited to take initiatives to enhance the co-operation and co-ordination on blue whiting research.

### 1.2 List of participants

| Alexander Krysov | Russia |
| :--- | :--- |
| Edda Johannesen | Norway |
| Asta Gudmundsdottir (Chair) | Iceland |
| Brian S. Nakashima | Canada |
| John Boyd | Ireland |
| Dimitri Vasilyev | Russia |
| Frans van Beek | Netherlands |
| Harald Gjösæter | Norway |
| Hjalmar Vilhjalmsson | Iceland |
| Ingolf Röttingen | Norway |
| Jan Arge Jacobsen | Faroe Islands |
| Jens Chr. Holst | Norway |
| Manuel Meixide | Spain |
| Mikko Heino | Norway |
| Nikolay Timoshenko | Russia |
| Sergei Belikov | Russia |
| Sigurd Tjelmeland | Norway |
| Sveinn Sveinbjörnsson | Iceland |
| Webjörn Melle | Norway |

### 1.3 Non-standard assessment methods

### 1.3.1 AMCI

The assessment model AMCI (Assessment Model Combining Information from various sources), version 2.1, was described in the Working Group report in 2002. For assessments in 2003 AMCI version 2.2 (May 2002) has been used. This version is essentially an updated version of AMCI 2.1 where some known problems have been solved but without important changes in functionality. An updated manual was available for the Working Group. The Working Group on Methods on Fish Stock Assessments explored and evaluated AMCI 2.2, together with ISVPA and an array of other assessment models in their meeting in early 2003 (ICES 2003/D:03). The report of that Working Group can be consulted for more details on AMCI.

### 1.3.2 SeaStar

The assessment model SeaStar (Stock Estimation with Adjustable Survey observation model and TAg return data) was described in the Working Group report in 2002 and has been used for assessing the Norwegian spring-spawning stock in recent years. An updated manual was available for the Working Group. This year the method for analysing young fish not yet present in the main tuning data was changed from a regression approach to tuning in a separate step using survey and 0 -group data from the Barents Sea.

### 1.3.3 ISVPA

The ISVPA is described in last years WGNPBW report (ICES 2002/ACFM:19). Some changes have been made to the model, and these are described in the report from the meeting of the Methods Working Group in January 2003 (ICES 2003/D:03).

### 1.4 Special requests concerning Norwegian spring-spawning herring from coastal states

### 1.4.1 Comparison of the SeaStar and ISVPA models

The special request from the coastal states is given in Section 1.1. An evaluation and comparison of the ISVPA and SeaStar and their applicability to assess the stock of Norwegian spring-spawning herring is given in Section 3.5.1 and 3.5.2.

### 1.4.2 Review of TAC advice for 2003

SeaStar assessment: The SeaStar assessment this year gives a similar perception of the spawning stock size for 2003 as last year. In the prognosis made in 2002 the assessment of the spawning stock for 2003 was 6.1 million $t$. This year the estimate of the spawning stock in 2003 is 5.8 million t. There has, however, been an increase in the estimate of the 1998 year class from 9.9 billion individuals as 4 years old to 11.5 billion in the 2003 estimate. A catch for 2003 corresponding to $\mathrm{F}=0.125$ using the same exploitation pattern as used in the TAC advice and based on the 2003 assessment of the stock in 2003 is 765000 tonnes, compared to 710000 tonnes based on the 2002 prognosis. The TAC advice for 2004 of 850000 t is based on a catch in 2003 of 710000 t . If the catch in 2003 is increased the TAC advice for 2004 should be reduced correspondingly.

ISVPA assessment: Norwegian spring-spawning herring stock assessment by means of ISVPA, made at the WG, confirmed the tendency of an increase in spawning stock biomass, of which the first indications were shown in last year's assessment by means of the model. Results of the ISVPA assessment indicate the possibility of considering an increase in the TAC level recommended by ACFM for 2003 from 710000 tonnes to 805000 tonnes, corresponding to $\mathrm{F}=0.125$.

### 1.5 Special requests concerning blue whiting

### 1.5.1 Medium-term projections

The Northern Pelagic and Blue Whiting Working Group did not carry out medium-term projections. In general, medium term projections in the most recent period are largely defined by the presently assumed stock size, recruitment and exploitation level, while the period further in the future is mostly affected by assumptions on recruitment dynamics and exploitation level. This year, the WG could not reach an agreement on the present situation in this stock, since two different models gave different answers. The Working Group considered that, given the uncertain present situation of the stock and its exploitation, and the poor knowledge on stock dynamics, medium-term projections for blue whiting would be predominantly based on assumptions and are not very instructive. However, the WG made medium-term projections during its 2002 meeting.

### 1.5.2 Harvest control rules

The Northern Pelagic and Blue Whiting Working Group did not carry out medium-term projections, which are the basis for the evaluation of harvest control rules. However, the WG made medium-term projections during its 2002 meeting, and, based on a similar request in that year, suggested and evaluated various harvest control rules. These should, in general, still be valid.

The Working Group considers that evaluation of harvest control rules for blue whiting would currently be premature. Evaluation of harvest control rules is possible if, and only if management objectives are well specified. The current management objectives for blue whiting are formulated within the framework of precautionary reference points, although this has had no influence on actual management. The reference points for blue whiting were derived in 1998 but have not yet been revised, despite the new dynamic regime displayed both by the stock and fisheries during the later years. Furthermore, the Working Group was not aware whether the evaluation should take into account management objectives beyond those implied by the precautionary reference points, involving, e.g., stability of landings.

### 1.5.3 Description of the fisheries

In a special request to ICES for scientific advice for 2004, NEAFC asked the following: "Under the MoU, ICES is requested to describe fisheries, gear, fleets and developments in fleets involved in fisheries for stocks for which NEAFC recommends management measures. The information is requested with the same level of detail as given for pelagic fisheries for Sebastes mentella in the Irminger Sea in the Report of the North Western Working Group."

The WG has consulted the proposed NWWG report and have decided to add a summary section of the "development of the fisheries" in sec. 6.2 in next years report. The data on the fisheries provided to this year's WG were not detailed enough to fully answer the request. However, in sec. 6.2 the spatial and temporal distribution of the catches of blue whiting in 2002 are given as catch by quarter and ICES rectangles (Figure 6.2.1) and for the whole year as catch by ICES rectangles (Figure 6.2.2).

In order to describe the developments in the fleets involved in the fisheries for blue whiting, Norwegian springspawning herring and capelin, the WG recommends that WG members bring data to enumerate the number, capacity and effort of vessels taking part in the fisheries.

### 1.5.4 Evaluation of the assessment models ISVPA and AMCI used for blue whiting

This request was addressed and answered by the ICES Methods Working Group during their meeting in January 2003 (ICES 2003/D:03).

### 1.5.5 Lack of coordination on research on blue whiting

The Northern Pelagic and Blue Whiting Working Group discussed this problem, and put forward the following recommendations:

Several surveys on blue whiting are presently going on. It is recommended that a coordinated survey be organised covering the main spawning grounds of blue whiting. Other countries than those presently taking part in these surveys should be invited to take part.

It is recommended that information from existing surveys in which blue whiting are caught is made available to the Working Group. In particular, information from PGSPFN-coordinated surveys should be made available and analysed for information on abundance on incoming year classes.

It is suggested that the coordination could be taken care of by an extended ICES PGSPFN.

The Working Group also acknowledged that the Nordic Council of Ministers have provided funds for the "Nordic Blue Whiting Network", which has already taken actions towards more coordinated research on blue whiting. The report of the network was made available to the Working Group (Heino et al., WD b).

### 1.5.6 Information about species composition in the industrial fisheries

Only information on species composition in catches from Norwegian mixed industrial fishery was made available to the Working Group (Heino et al., WD a). This fishery targets both blue whiting and Norway pout, and the landings consist of mixtures of these two species. In 2000-2002, the average annual landings from the fishery were 109000 t . Of this amount, an estimated proportion of $58 \%$ was blue whiting, and $17 \%$ was Norway pout. The remaining proportion, $15 \%$ or about 16000 t , represents a wide range of fish and invertebrates. The six most important bycatch species (in terms of landed catch) are saithe, herring, haddock, horse mackerel, whiting and mackerel, each of which represents an annual catch of at least 1000 t in this fishery. Of these species, mostly individuals in the length range $25-40 \mathrm{~cm}$ were captured, with herring and mackerel being often somewhat smaller and saithe larger. This suggests that the bycatch of these species often consists of immature individuals.

The Working Group does not have expertise for judging the significance of bycatch mortality on species other than blue whiting. The Working Group therefore recommends that working groups responsible for the assessments of the species concerned carry the analyses further.

### 2.1 Climate considerations in the Barents Sea

### 2.1.1 Hydrography and ice conditions

The Barents Sea is characterised by large year-to-year fluctuations in heat content and ice coverage caused by variations in the influx of Atlantic water from the Norwegian Sea (Figure 2.1.1.1). Temperatures in the Barents Sea have been relatively high during most of the 1990s, with a continuous warm period from 1989-1995. During 1996-1997, the temperature was just below the long-term average before turning warm again at the end of the decade. The 1990s was the third warmest decade in the $20^{\text {th }}$ century (Ingvaldsen et al., in press).

In January 2002 the temperature was just above the long-term average in the whole Barents Sea, but from April the temperature increased rapidly. In the Fugløya-Bjørnøya section (Figure 2.1.1.2 and Figure 2.1.1.3) the temperature in June was $1^{\circ} \mathrm{C}$ above average, which is the highest observed value since the start of measurements in 1977. In the Kola section the maximum temperature was $0.8^{\circ} \mathrm{C}$ above average in August/September, which was $0.1-0.2^{\circ} \mathrm{C}$ below the maximum for the period 1921-1999. The temperature decreased slightly until October, followed by a rapid decrease towards the average in December. In January 2003, the temperature was at the long-term average (Asplin and Dahl, 2003; Stiansen et al., WD). The situation was similar in the whole Barents Sea.

The variability in the ice coverage is closely linked to the temperature of the inflowing Atlantic water. The ice has a relatively short response time on temperature changes in the ocean, but usually the sea ice distribution in the eastern Barents Sea responds a bit later than in the western part. 2001 had the highest ice index recorded since 1970, which means very little ice. 2002 had the second highest ice index. During the winter of 2002 there was about the same ice conditions as the year before, but the ice melt during summer was quite high. The winter of 2003 will have more ice than 2002, but the ice index is still expected to be higher than average for the whole year.

### 2.1.2 Inflow of Atlantic water

Transport of Atlantic water to the Barents Sea has been measured since August 1997. The flow of Atlantic water is very variable. Most of the time there is a net inflow of Atlantic water to the Barents Sea, but in some periods large outflows are observed. Large outflows occurred in April in 1998 and 1999, and in 2000 there were two periods with strong outflows, one in January and one in June. In January and March 2002 there were two peaks of high inflow into the Barents Sea. The intensity of the flow was reduced during spring and summer. In October 2002 there was a peak of weak outflow. Results from a wind driven model shows similar results. The inflow from the model during the first two months was stronger than average. The rest of the year the model showed average inflows, except for the last two months when the flow was reduced.

### 2.1.3 Predicting Barents Sea temperature

Prediction of Barents Sea temperature is complicated since the variation is governed by processes of external and local origin that operate on different time scales (Stiansen et al, WD). The volume flux and temperature of inflowing Atlantic water masses, as well as heat exchange with the atmosphere, is important in determining the temperature of the Barents Sea. Thus, both slowly moving advective propagation and rapid barotropic responses due to large-scale changes in air pressure must be considered. The major changes in Barents Sea climate take place during the winter months. The variability in the amount of heat flowing in with Atlantic water masses from the south is particularly high during this season. Furthermore, variability in low-pressure passages and cloud cover has a strong influence on the winter atmosphere-ocean heat exchange.

This seasonal difference is reflected in the merit of simple six-month forecasts of Kola-section temperature based on linear regression models. The tendency is that persistence across the spring and summer months is higher than for other seasons, allowing for reasonably reliable forecasts from spring until autumn. Data available until February 2003 allow for a six-month forecast for August 2003. The value for February 2003 of $3.3{ }^{\circ} \mathrm{C}$ is inserted into the equation $\mathrm{T}_{\text {August }}=$ $2.37+0.67 * \mathrm{~T}_{\text {February }}$, statistically derived from data for the years 1921-1997 (Stiansen et al, WD). This gives an objective temperature forecast for August 2003 of $4.58^{\circ} \mathrm{C}$. This will be slightly below the 1921-1999 mean of $4.67^{\circ} \mathrm{C}$. We conclude that summer sea temperatures in the southern Barents Sea are expected to lie around the long-term mean.

## Conclusions:

- 2002 was warmer than average. The temperature in the beginning of the year was just above average, followed by an extremely hot summer, while the temperature decreased below the average at the end of the year.
- The inflow of Atlantic water was normal for most of 2002, except for a higher inflow at the beginning of the year.
- The temperature in 2003 is expected to be lower than in 2002, and will be close to the long-term mean in most of the Barents Sea.


### 2.2 Zooplankton

The standing stock of zooplankton has been monitored by IMR in the Barents Sea from the early 1980s in connection with the joint Norwegian/Russian 0-group and capelin surveys during August-October. At this time of the year most of the production has taken place and the zooplankton biomass can be expressed as the overwintering population of zooplankton. Plankton samples were obtained using WP2 and the MOCNESS (Multiple Opening Closing Net and Environmental Sensing System) plankton net. In 2002 PINRO also joined in the collection of samples of zooplankton during August/October. Plankton samples in Russian surveys are collected using the Juday net.

The mean biomass $\left(\mathrm{gm}^{-2}\right)$ values from 1988 to the present are estimated for the 7 different areas in the Barents Sea. There was a marked increase in zooplankton biomass during the period 1991-1994. The highest biomass values were observed in 1994 when the capelin stock was at an extremely low level. Though the biomass has decreased from 1994 to the present, the average biomass values during 1995 to 2002 are still higher than in the 1988-1992 period. In 2002 the zooplankton biomass was average, with a slight increase from 2001 to 2002.

Figure 2.2.1 shows the total biomass of zooplankton together with capelin stock size (million tonnes). A commonly observed inverse relationship between capelin stock size and zooplankton biomass can be seen from Figure 2.2.1, indicating that capelin exercise a strong feedback control on the system through its predation pressure on zooplankton.

## Conclusion:

- A moderately overwintering zooplankton biomass in 2002 above the average will create the basis for average zooplankton production in 2003 and feeding conditions for capelin, as well as for other pelagic fish and juvenile demersal species in the Barents Sea.


### 2.3 Trophic interactions

### 2.3.1 Predicting capelin biomass

Capelin is the most important prey species for Northeast Arctic cod, and the development of the capelin stock may have a strong effect on growth and maturation of cod, as well as cod cannibalism.

The biomass of capelin $(1+)$ decreased from 3.6 million tonnes in 2001 to 2.2 million tonnes in 2002 (ICES 2002/ACFM:19). This is lower than the prediction for 2002 made by AFWG last year ( 3.4 million tonnes). The prediction method used in ICES (2002/ACFM:19), which is essentially the same as the one used previously, predicted the $2+$ capelin biomass to be 1.40 million tonnes in October 2003 and the biomass of 1 -year-old capelin at the same time to be 0.59 million tonnes, giving a total of 1.99 million tonnes. Of this, 1.17 million tonnes are predicted to be mature capelin. The stock history for capelin from 1984 onwards is given in Table 2.3.1.1 together with the estimated biomass of capelin removed from the stock by natural mortality.

A 1-year prognosis has been presented to AFWG since 1999. A review of the prognoses made during this period is given in Table 2.3.1.2. The prognoses seem to be overestimates in most cases. The prediction methodology is still under development. WGNPBW has been requested by AFWG to provide a review of how the present prognosis method would have performed when run on historical data. Also, the prediction should be given with uncertainty. This will be done during the capelin assessment sub-group meeting after the joint capelin survey in the autumn.

### 2.3.2 Predation by cod

The consumption by cod of various prey species for the period 1984-2002 is given in Table 2.3.2.1, using the same method as described by Bogstad and Mehl (1997).

As usual, capelin was the most important prey for cod. However, the consumption of capelin by cod decreased markedly from 2001 to 2002. This may be related to the decrease in the capelin stock. The consumption by cod of other fish species (herring, polar cod, cod, haddock and blue whiting) increased from 2001 to 2002 . The consumption of blue
whiting increased to 277000 tonnes, the highest value in the 19-year time-series. The consumption of shrimp, krill and amphipods decreased from 2001 to 2002.

Dolgov (WD to AFWG, Table 2.3.2.2) also calculated the consumption by cod based on the same data, using a somewhat different methodology. The consumption by prey species from the two methods for 2002 is similar. The main difference is that the calculations in Table 2.3.2.1 give a decrease in the consumption of capelin from 2001 to 2002, while the calculations in Table 2.3.2.2 show an increase. Also, there are notable differences in the number-at-age of cod and haddock consumed by cod. It should be noted that the calculations in Table 2.3.2.1 are based on the number-at-age of cod from the VPA given in this year's report, while the calculations in Table 2.3.2.2 are based on the VPA from the 2002 AFWG meeting.

The annual consumption for each age group of $\operatorname{cod}(\mathrm{kg} /$ year $)$, based on the consumption calculations shown in Tables 2.3.2.1 and 2.3.2.2 are given in Tables 2.3.2.3 and 2.3.2.4, respectively. Table 2.3.2.3 shows that the consumption per cod decreased from 2001 to 2002 for age 3 and older fish. The consumption per cod in 2002 was close to the 1998 level, but lower than in the period 1999-2001. Such a trend in the consumption per cod is not found in Table 2.3.2.4. The calculations by Dolgov (WD to AFWG) generally give a lower consumption per cod for age 1-4 and a higher consumption per cod for age 6+ compared to the calculations using the method described by Bogstad and Mehl (1997). The discrepancies in consumption per cod by age group are much larger than the discrepancies in total consumption by the cod stock.

The consumption estimates in Tables 2.3.2.1 and 2.3.2.2 do not include the consumption by mature cod in the period when it is outside the Barents Sea (assumed to be 3 months during the first half of the year). During this period it may consume significant amounts of adult herring (Bogstad and Mehl, 1997).

Johansen et al. (2002) describe a new method for calculating the consumption by cod, and used it to calculate the consumption of herring by cod in the period 1992-1997. Their consumption estimates are comparable to the estimates given in Table 2.3.2.2, except for 1994, when they obtained a much higher estimate ( 494 vs .147 thousand tonnes).

As in previous years, the consumption of cod and haddock by cod was calculated using the method described by Bogstad and Mehl (1997). It is important to agree on a joint methodology for consumption calculations.

### 2.3.3 Predation by other fish species

Dolgov et al. (WD to AFWG, AFWG 2002) investigated the diet of blue whiting in the Barents Sea in the period 19982001. They concluded that predation by blue whiting will not have a significant impact on the recruitment of cod, haddock and redfish. However, food competition between blue whiting and juveniles of other commercial fish stocks due to blue whiting grazing zooplankton in the areas of larval drift may occur. The diet of saithe in the period 19982001 was investigated by Dolgov (WD to AFWG).The diet of saithe $>40 \mathrm{~cm}$ is dominated by capelin, with herring and euphausiids being next in order of importance. In some areas there are significant amounts of blue whiting and haddock juveniles. For saithe $<40 \mathrm{~cm}$, the diet is dominated by euphausiids.

### 2.3.4 Predation by mammals

The consumption by minke whales (Folkow et al,. 2000) and by harp seals (Nilssen et al., 2000) is given in Table 2.3.4.1. These consumption estimates are based on stock size estimates of 85000 minke whales in the Barents Sea and Norwegian coastal waters (Schweder et al., 1997) and of 2223000 harp seals in the Barents Sea (ICES 1999/ACFM:7). The consumption by harp seal is calculated both for situations with high and low capelin stock, while the consumption by minke whale is calculated for a situation with a high herring stock and a low capelin stock. It is worth noting that the abundance estimate of harp seals was revised considerably upwards in 1998 (ICES 1999/ACFM:7), which also increased estimates of the consumption by harp seals correspondingly. Food consumption by harp seals and minke whales combined is at about the same level as the food consumption by cod, and the predation by these two species needs to be considered when calculating the mortality of capelin and young herring in the Barents Sea.

In the period 1992-1999, the mean annual consumption of immature herring by minke whales in the southern Barents Sea varied considerably ( $640 \mathrm{t}-118000 \mathrm{t}$ ) (Lindstrøm et al., 2002). The major part of the consumed herring belonged to the strong 1991 and 1992 year classes and there was a substantial reduction in the dietary importance of herring to whales after 1995, when a major part of both the 1991 and 1992 year classes migrated out of the Barents Sea. In 19921997, minke whales may have consumed 230000 t and 74000 t , corresponding to 14.6 billion and 2.8 billion individuals of the herring year classes of 1991 and 1992, respectively. The dietary importance of herring to whales appeared to increase non-linearly with herring abundance.

Conclusions:

- The capelin biomass in 2003 is expected to be approximately the same as in 2002, which suggests that the decline observed in recent years has been halted.
- The consumption of capelin by cod decreased from 2001 and 2002, according to Norwegian consumption calculations, but increased according to the Russian calculations.
- The consumption of other fish species by cod increased from 2001 to 2002, while the consumption of shrimp, amphipods and krill decreased from 2001 to 2002.
- The consumption per cod decreased from 2001 to 2002 according to Norwegian calculations, while Russian calculations showed a stable consumption by cod.
- There was also some decrease in the consumption of capelin by cod.


### 2.4 Norwegian Sea

### 2.4.1 Hydrography and climate

The Nordic Seas during the last decades have been characterized by increased input of Arctic waters. The Arctic waters to the Norwegian Sea are mainly carried by the East Icelandic Current and also to some extent by the Jan Mayen Current. During periods of increased Arctic water input, the western extension of Atlantic water is moved eastward. As a result, over the last 25 years the southern and western Norwegian Sea has become colder and fresher while the eastern Norwegian Sea is warmed. Atmospheric forcing drives this trend. Since the mid 1960's the North Atlantic Oscillation index (NAO) has increased (Figure 2.4.1.1). NAO as it is used here is the normalised air pressure difference at sea level between Lisbon, Portugal and Reykjavik, Iceland and is an indicator of the strength of the westerly winds into the Norwegian Sea. A high NAO index (i.e. stronger westerly winds) will force Atlantic and Arctic waters more eastward.

The Institute of Marine Research, Norway, has measured temperature and salinity in three standard sections in the Norwegian Sea almost regularly since 1978 (Figure 2.1.1.3). The sections are 1) the Svinøy section which runs NW from $62.37^{\circ} \mathrm{N}$ at the Norwegian coast, 2) the Gimsøy section which also runs NW from the Lofoten Islands and 3) the Sørkapp section which is a zonal section at $76.33^{\circ} \mathrm{N}$ just south of Svalbard.

Figure 2.4.1.2 shows the development in summer (July-August) temperature and salinity in the sections from south to north in the Norwegian Sea (Melle et al., WD). During the last 6 years the temperature and salinity in the Svinøy section have been above the long-term mean, while they were about average in the Gimsøy and Sørkapp sections. In 2002 there was a large increase in both temperature and salinity in the Svinøy section. The temperature was then the largest value in the time-series, about $1.3^{\circ} \mathrm{C}$ above the normal, while the salinity was the next largest, 0.07 above the normal. Only in 1983 was the salinity higher. This increase in temperature and salinity was not seen further north in the Sørkapp section. Unfortunately there were no observations in Gimsøy section that summer.

Figure 2.4.1.3 shows time-series of temperature and salinity during the spring in the Svinøy and Gimsøy sections from 1978 to 2003. The values are calculated using the same procedure as mentioned above. The low salinities in 1978 and 1979 are a result of the Great Salinity Anomaly during the 1970s. In 1994 a large salinity anomaly comparable with the anomaly in 1978 and 1979 was seen in the Svinøy section. The temperature was also a minimum that year. The 1994 anomaly was a result of increased influence of Arctic water from the East Icelandic Current. In 2002 the salinity and temperature increased considerably in the Svinøy section to the highest observed value for the time-series. The condition in 2003 remained approximately unchanged compared to 2002. In the Gimsøy section there was instead a reduction in temperature and salinity for 2002, but both increased again in 2003. The salinity was then the highest since 1985.

The area of Atlantic water (defined with $\mathrm{S}>35.0$ ) in the Svinøy-section has been calculated. The mean temperature within the limited area has also been calculated, and the results for both spring and summer are shown in Figure 2.4.1.4. There are considerable variations both in the area of Atlantic water distribution and its temperature. The distribution area of Atlantic water has decreased since the beginning of the 1980s, while the temperature has shown a steady increase. Since 1978 the Atlantic water has been about $0.5^{\circ} \mathrm{C}$ warmer. During the years 1992-1995 the area was much lower than average for both seasons. In 1997-1999 there was a warm period followed by a substantial drop in temperature in 2000. Then in 2002 the temperature increased considerably and had the highest values in both timeseries. The temperature in 2002 was $0.7^{\circ} \mathrm{C}$ higher than the long-term mean for both spring and summer. While the temperature increased significantly, the area of Atlantic water in 2002 was close to normal.

Conclusions:

- Temperature and salinity in the Svinøy section were the highest ever for 2002 and remained high also during spring 2003.
- The Gimsøy and Sørkapp sections did not have as high values of temperature and salinity in 2002, but the salinity for spring 2003 in the Gimsøy section was the highest measured since 1985.
- The averaged temperature of the Atlantic Water in the Svinøy section has increased approximately $0.5^{\circ} \mathrm{C}$ since 1978.
- The winter NAO in 2003 is expected to be close to normal.


### 2.4.2 Phytoplankton

The development of phytoplankton in the Atlantic water is closely related to the increase of incoming solar irradiance during March and to the development of stratification in the upper mixed layer due to warming. The Institute of Marine Research, Norway, started in 1990 a long-term study of the mechanisms controlling the development of phytoplankton at Ocean Weather Station Mike situated at $66^{\circ} \mathrm{N}, 2^{\circ} \mathrm{E}$.

Figure 2.4.2.1 shows the development of the phytoplankton bloom for 2002 (Melle et al., WD). In previous years there has been a marked difference in the time when the spring bloom reached its maximum. In 1997 the spring bloom reached its maximum 20 May (day of the year 140), in 1998 about one month earlier 18 April (day of the year 108). The timing of the bloom in 1999 was similar to that in 1998, but did not show the same high maximum in chlorophyll. This may be related to the weekly measurements in 1999, as opposed to daily measurements in 1997 and 1998. On the other hand, weekly measurements prior to 1997 have revealed pronounced maxima in chlorophyll. The reason for the low algal biomass in 1999 may have been early and strong grazing from a large over-wintering zooplankton stock. In all these years a strong peak has characterized the bloom. The situation in 2001 was different from previous years. First, the spring bloom started somewhat later (first week of May) compared to 1998 and 1999 and was followed by relatively moderate chlorophyll concentrations culminating with a major peak in the first week of June. Also a distinct early autumn bloom was observed in the middle of August. In 2002 the springbloom started to develop in the middle of April reaching its maximum at the end of April, resulting in one of the earliest blooms, second only to the bloom in 1998. The 2002 bloom also maintained relatively high chlorophyll concentrations for about three weeks after the peak. The development of phytoplankton prior to the spring bloom may be separated into two phases. The first phase, from day 1 to about day 50 , is characterised by extremely low phytoplankton biomass expressed as chlorophyll $a$. This is the winter season during which phytoplankton growth is mainly limited by the low incoming irradiance typical of this period. The second phase, from about day 50 to day 100 , is characterised by a gradual increase of phytoplankton biomass but without reaching bloom conditions. This is the pre-bloom phase during which the increase in biomass is related to the increase in incoming irradiance, and the lack of a bloom is due to the deep upper mixed layer still being present at this time.

Figure 2.4.2.2 shows the extension in time for these two phases and the timing of the spring bloom for the period 19912002. In a "normal" year the winter season extends to about 2 March. The pre-bloom phase extends on average from 2 March to 16 April. The spring bloom starts normally on 16 April and reaches its maximum on 21 May, but the year-toyear variations are much larger than those of the previous phases. From 1991 to 1995 the trend was towards earlier spring blooms. This trend was broken in 1996, and thereafter the year-to-year variability in the timing of the bloom has been greater.

## Conclusions:

- The phytoplankton bloom in 2002 developed quite early, second only to the 1998 bloom.
- Chlorophyll $a$ concentrations first peaked in late April 2002 and were maintained at relatively high levels until the third week of May. This could have been the result of a delay in grazing pressure.
- During summer and early autumn several peaks of relatively high chlorophyll $a$ concentrations were observed, indicating a strong variability in minor blooms.


### 2.4.3 Zooplankton

Zooplankton biomass distribution in the Norwegian Sea has been mapped annually in May since 1995 and in July since 1994. Zooplankton samples for biomass estimation were collected by vertical net hauls (WP2) or oblique net hauls (MOCNESS). In the present report zooplankton samples from the upper 200 m are analysed. Total zooplankton biomass ( g dry weight $\mathrm{m}^{-2}$ ) in May was averaged over sampling stations within three water masses - Atlantic water (defined by salinity $>35$ at $20-\mathrm{m}$ depths), Arctic water (salinity $<35$, west of $1.4^{\circ} \mathrm{E}$ ) and Coastal water (salinity $<35$, east of $1.4^{\circ} \mathrm{E}$ )
(Figure 2.4.3.1). In Atlantic and Arctic water masses zooplankton biomass decreased to a minimum in 1997 (Melle et al., WD). Thereafter zooplankton biomass increased again and has remained relatively high except for a temporary reduction in 2001. Due to reduced cruise time the Arctic water mass was not sampled in 2001. For the first time in 2002, the biomass in Atlantic water equalled the biomass in Arctic water. In the Coastal water masses, which includes the Norwegian continental shelf and slope waters influenced by Norwegian coastal water, the trend was different with a general increase towards a maximum in 1998 and a decrease in the following years. Biomass increased again in 2002, and reached the highest value for the time-series.

In July the total zooplankton biomass ( g dry weight $\mathrm{m}^{-2}$ ) in the upper 200 m was calculated by integrating biomass at sampling stations within a selected area in the central and eastern Norwegian Sea. There is no obvious trend in the zooplankton biomass in July since 1994 (Figure 2.4.3.2).

Conclusions:

- Average zooplankton biomass in Atlantic water masses of the Norwegian Sea in May 2002 was close to the mean for the time-series.
- Zooplankton biomass in July 2002 was higher than in 2001.


### 2.4.4 Herring growth and food availability

Individual growth of the Norwegian spring-spawning herring, as measured by condition or length-specific weight after the summer feeding period in the Norwegian Sea, has been characterised by large fluctuations during the 1990s (Figure 2.4.4.1). During 1991 and 1993 individual condition was good, but from 1994 on the condition of the herring started to decline and by 1997 it reached the lowest level during the 1990's. The level observed in 1997 corresponds with the absolute long-term low level observed during the period 1935 - 1994 (Holst 1996). Following a recovery in 1998 and 1999, the condition of the herring decreased again. In 2001 and 2002 the condition remained at a low level (Melle et al., WD).

Since 1995, when the large-scale migration pattern of herring has been mapped during two annual cruises, May and July-August, the herring have been feeding most heavily in Atlantic water, and the herring condition index obtained after the feeding period in the Norwegian Sea is related to the average zooplankton biomass of Atlantic water (Figure 2.4.4.2). To improve this relationship herring feeding areas should be defined more precisely, because large variations in herring migration routes and in zooplankton distribution have been observed over the years.

Conclusions:

- Herring condition increased from 2001 to 2002.
- There is a weak relationship between zooplankton biomass in May and herring condition in the autumn during the years 1995-2002.


### 2.4.5 Predictions for zooplankton biomass and herring feeding conditions

A factor possibly governing zooplankton biomass is the size of the zooplankton spawning stock, or the size of the overwintering population (Melle et al., WD). Zooplankton biomass in July represents the mixed population of zooplankton species at the start of the over-wintering period. A linear regression of the biomass in July on the biomass in May the following year explains $\sim 63 \%$ of the total variation (Figure 2.4.5.1). Average biomass in July 2002 suggests that zooplankton biomass in May 2003 will be close to the average as well (Figure 2.4.5.1). However, the time-series is short, the variability is larg, e and there is no trend in the July zooplankton biomass that could be related to the trend observed in the May data. Thus, this time-series should be expanded before it is used for prediction.

The North Atlantic Oscillation index (NAO), is a proxy for the strength and duration of southwesterly winds, and is correlated with the inflow of Atlantic water to the Norwegian Sea. In the Norwegian Sea the winter NAO (average of the indices from December to March) is correlated with the zooplankton biomass in May the following year (Melle et al,. 2002). This relationship was explored in more detail in the present report, an exercise that we hoped would reveal relationships that reflect the causal relationship between climate and zooplankton production more closely. The NAO index was averaged over four two-month periods; January-February, February-March, March-April and April-May. Then biomass in May was correlated with the mean NAO for the two-month periods the previous year. The strongest correlation was found between biomass in May and the average NAO for the March-April period the previous year (Figure 2.4.5.2). March-April is the period when the primary production in the Norwegian Sea is initiated and the major reproductive period for many important zooplankton species such as Calanus finmarchicus and krill. The one-year lag
in the relationship may be because in May we mainly measure the size of the previous year's overwintering stock, that is the previous year's production and the present year's spawning stock. The biomass for May 2003 is predicted at 12.98 g dry weight $\mathrm{m}^{-2}$, based on the NAO for 2002. When the NAO for April 2003 is available the biomass for May 2004 can then be predicted.

Biomass (yr2) $=2.74 * N A O$ yrl $1+11.61$
$R^{2}=0.78, P=0.004$ The time-series for the herring condition index has been calculated for the period from 1991 to 2001. A correlation analysis of herring condition on the two-month average of the NAO indices showed that the relationship was strongest between herring condition and the NAO during the March-April period (Figure 2.4.5.3). The herring condition index for 2003, based on the NAO for 2002, is predicted to be 0.833 . This is similar to the condition in 2002 (0.808). When the NAO for April 2003 is available, the condition for 2004 can then be predicted.

Condition (yr2) $=0.022 * N A O$ yr1 $1+0.822$ (2)
$R^{2}=0.66, P=0.007$ Conclusions:

- A direct, but weak, relationship between zooplankton biomass in July and the zooplankton biomass in May the following year is suggested by the time-series from 1994 to 2002.
- The average NAO for March-April the previous year is directly related to zooplankton biomass in May and herring condition in the autumn.
- The March-April NAO index for 2002 predicts zooplankton biomass at $12.98 \mathrm{~g} \mathrm{~m}^{-2}$ in May 2003 and the herring condition index at 0.833 in the autumn 2003.


### 2.5 Icelandic Waters

### 2.5.1 Hydrography and climate

Due to the proximity of the oceanic Polar Front in the northern North Atlantic, hydrographic conditions in the sea north of Iceland are highly variable. Changes in intensity of the influx of Atlantic water and/or the variable admixture of polar water to the surface layers north of Iceland may lead to marked fluctuations in temperatures and salinities, both in space and time. Off the south and west coasts, where Atlantic water predominates, fluctuations are much smaller.

Climatic conditions in the North Atlantic improved greatly around 1920 and remained good until the mid-1960s, when they deteriorated suddenly. In the area north and east of Iceland temperature and salinity declined sharply in 1965 and these severely cold conditions lasted until 1971. After that, climatic conditions of the area north and east of Iceland improved again, but were variable and warm years have alternated with cold years (Figure 2.5.1.1).

In the latter half of 1997, there was a pronounced increase in the intensity of the Irminger Current south and west of Iceland, resulting in temperatures $\left(6-8^{\circ} \mathrm{C}\right)$ and salinities (35.0-35.2) similar to those recorded in these waters in the 1950s and the early 1960s. There were no signs of a reduction of this flow of warm water off south and West-Iceland in February or May 2002. The inflow of Atlantic water to the north Icelandic area was not quite so pronounced in these months as in 2001 (Figure 2.5.1.1) and the cold East Icelandic Current was not as far offshore as in 2001. There were, however, some indications of an increased inflow of Atlantic water off the western north coast of Iceland in May 2002. By August the flow of Atlantic water onto the north Icelandic shelf had reached a level similar to that recorded in previous years and these warm conditions were found to prevail in November 2002 and February/March 2002.

### 2.5.2 Zooplankton

In the area north of Iceland, zooplankton biomass is significantly higher during years with a strong inflow of Atlantic water than in years when Atlantic inflow is weak and lower salinity in the surface layer slows or prevents vertical mixing (Figure 2.5.1.1). In spring 2002, the zooplankton biomass for the whole Icelandic area was near the long-term average. Off West- and East-Iceland the zooplankton biomass was well above average, but slightly below the long-term mean south of Iceland. The relatively low zooplankton biomass off the central north coast in spring 2002 is in line with reduced Atlantic influence as compared to the most recent years. As in the last few years, the biomass of cold-water species in the deep waters off Northeast- and East-Iceland was very high.

As mentioned above, a continued strong inflow of Atlantic water to the north Icelandic area was observed both during the November 2002 and February/March 2003 surveys. This indicates that zooplankton biomass will be above average in north Icelandic waters in spring and summer 2003.

### 2.6 Hydrography of the waters west of the British Isles

The hydrography of the waters west of the British Isles is described in a cruise report by Heino et al. (WDc). The horizontal distribution of temperature at 10 and 400 meters depths are shown in Figures 2.6.1 and 2.6.2, respectively. The maps are based on data collected on board "Johan Hjort", and CTD data are kindly provided by the scientists on board the Russian ships "Smolensk" and "Atlantniro", who were running simultaneous surveys in the area. The cooperation has given a much better horizontal coverage of the area.

The Wyville-Thompson ridge $\left(\sim 60^{\circ} \mathrm{N}\right)$ divides the survey area into two very different hydrographic regimes. South of the Wyville-Thompson ridge the vertical gradients in temperature are small. Temperatures at 1000 m are typically between 7 and $8{ }^{\circ} \mathrm{C}$, i.e., the vertical temperature decreases by only $2-3^{\circ} \mathrm{C}$ from the surface to $1000-\mathrm{m}$ depth (see Figure 2.6.3), and in the top 600 m the temperatures drop by only about $1^{\circ} \mathrm{C}$. In the Faroe-Shetland channel the situation is different with a strong thermocline around $500-\mathrm{m}$ depth separating a layer of warm saline Atlantic water overlying cold deep waters $\left(\sim-0.5^{\circ} \mathrm{C}\right)$ originating in the Norwegian Sea (see Figure 2.6.4, Faroe-Shetland section).

Also the horizontal gradients are generally very small in the area south of the Wyville-Thompson ridge, and in particular the north-south gradient is very small. In the Rockall Trough the temperature drops by less than $2^{\circ} \mathrm{C}$ from $50^{\circ} \mathrm{N}$ to $60^{\circ} \mathrm{N}$ both at $10-\mathrm{m}$ and $400-\mathrm{m}$ depths (Figures 2.6.1 and 2.6.2). Due to a northward flowing shelf edge current, warm high salinity ( $\mathrm{S}>35.45$ ) water penetrates far north in a narrow band along at the shelf edge, with the $10^{\circ} \mathrm{C}$ isotherm at $10-\mathrm{m}$ depth extending north into the Faeroes-Shetland channel (Figure 2.6.1). Visual inspection of the sections and horizontal temperature maps indicates that this year's temperatures are up to $0.5^{\circ} \mathrm{C}$ and salinities up to 0.05 higher than in 2002. The vertical section plot of temperature and salinity (Figure 2.6.4) shows that the Atlantic water occupies all the area above 500 m and the $0^{\circ} \mathrm{C}$ isotherm is depressed down to 700 m . The area occupied by the warm Atlantic water is larger and the maximum temperature is higher than previous years.

The high temperatures and salinities are confirmed by a study of the temperatures and salinities on all blue whiting cruises from 1983 through 2003. Since the hydrographic surveys have been dependent on the fishery surveys, the CTD stations have been distributed along the shelf edge and have in general not been in the same positions from year to year. In order to make time-series, the data were grouped in boxes with horizontal dimensions of $2^{\circ}$ latitude times $2^{\circ}$ longitude. For each year the mean temperature and salinity from 50 to 600 m of all the stations in deep water (depth $>600 \mathrm{~m}$ ) in each box was calculated. Some of the boxes had good coverage nearly every year, while others had many years missing. However, in general the same variation from year to year was seen in the boxes along the shelf edge south of the Wyville-Thompson ridge. The box with limits $52^{\circ}$ to $54^{\circ} \mathrm{N}$ and $16^{\circ}$ to $14^{\circ} \mathrm{W}$ had few gaps; the time-series of mean temperature and salinity for this box is shown in Figure 2.6.5. The pattern seen is that after some years with temperatures around $10.1^{\circ} \mathrm{C}$ in the 1980 s , it dropped to a minimum in $1994\left(\sim 9.8^{\circ} \mathrm{C}\right)$. After 1994 an increase in temperature is seen, and in 1998 temperature reaches a local maximum $\left(\sim 10.5^{\circ} \mathrm{C}\right)$ with the three following years a few tenths of a degree colder. 2002 is the warmest with $\sim 10.7^{\circ} \mathrm{C}$ and in 2003 the temperature was about the same as in 1998. A closer inspection shows that the decrease in temperature is caused by a lower temperature in the deep part of the layer, whereas in the upper part it is the same as last year. The vertical gradient within this layer was very small last year with a change in temperature of only about $0.7^{\circ} \mathrm{C}$ from 50 m to 600 m , but this year it dropped by $1.3^{\circ} \mathrm{C}$.

In the boxes along the continental shelf in the Rockall Trough a similar pattern as described above is seen in the timeseries, but the temperatures did not peak in 2002 and the temperatures in 2003 are higher or at least as high as in 2002. Thus in the Rockall Trough the temperatures in 2003 from 50 m to 600 m are the highest on record. In the northern part of the Rockall Trough the temperatures in the $50-600 \mathrm{~m}$ layer are typically about $0.5^{\circ} \mathrm{C}$ and salinity 0.05 higher than in 2002. In the shallow layer $50-150 \mathrm{~m}$, the temperatures are the highest on record for the whole area, and in the northern part the temperature is more than $0.5^{\circ} \mathrm{C}$ higher than last year.

The temperatures in the whole area are high in 2003, and except for the area to the west of Porcupine Bank, 2003 stands out as the warmest year in the observation period from 1983 to 2003. There is no clear linear trend, but the last five years are clearly warmer than the average of the whole period (1983-2003), and about $0.5^{\circ} \mathrm{C}$ above the first years in the period. Even though the increase is not as evident in the salinity curve, the high temperatures are typically associated with high salinities (Figure 2.6.5).

Table 2.3.1.1. Capelin stock history from 1984, and prognosis for capelin biomass in 2003. M output biomass is the estimated biomass of the capelin removed from the stock by natural mortality.
$\left.\begin{array}{|l|l|l|l|}\hline \text { Year } & \begin{array}{l}\text { Total stock number, } \\ \text { billions (Oct. 1) }\end{array} & \begin{array}{l}\text { Total stock biomass } \\ \text { in 1000 tonnes } \\ \text { (Oct. 1) }\end{array} & \begin{array}{l}\text { M output biomass } \\ \text { (MOB) during year }\end{array} \\ (1000 \text { tonnes) }\end{array}\right]$

* Estimate, includes the 2002 year class, whose size is estimated from a regression on an 0 -group index.

Table 2.3.1.2. Capelin one-year prognoses compared with survey estimates (in million tonnes).

| Year | Prognosis (1+ capelin biomass) <br> Available at AFWG in this year | Survey estimate (1+ capelin biomass) |
| :--- | :--- | :--- |
| 1999 | 4.0 | 2.8 |
| 2000 | 3.8 | 4.3 |
| 2001 | 4.1 | 3.6 |
| 2002 | 3.4 | 2.2 |


| Year | Other | Amphipods | Krill | Shrimp | Capelin | Herring | Polar cod Cod |  | Haddock | Redfish | G. halibut | Blue whiting | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 506 | 27 | 112 | 436 | 722 | 78 | 15 | 22 | 50 | 364 | 0 | 0 | - 2332 |
| 1985 | 1157 | 169 | 57 | 155 | 1619 | 183 | 3 | 32 | 47 | 225 | 0 | 1 | 3649 |
| 1986 | 665 | 1223 | 108 | 142 | 835 | 133 | 141 | 83 | 110 | 313 | 0 | 0 | - 3754 |
| 1987 | 680 | 1084 | 67 | 191 | 229 | 32 | 205 | 25 | 4 | 324 | 1 | 0 | 2843 |
| 1988 | 407 | 1236 | 317 | 129 | 339 | 8 | 92 | 9 | 3 | 223 | 0 | 4 | 42767 |
| 1989 | 725 | 800 | 241 | 132 | 580 | 3 | 32 | 8 | 10 | 232 | - | 0 | 2765 |
| 1990 | 1447 | 136 | 83 | 194 | 1593 | 7 | 7 | 19 | 15 | 243 | 0 | 85 | - 3829 |
| 1991 | 1076 | 65 | 75 | 188 | 2902 | 8 | 12 | 26 | 20 | 312 | 7 | 10 | 4702 |
| 1992 | 1014 | 102 | 157 | 373 | 2455 | 331 | 97 | 54 | 106 | 189 | 20 | 2 | 24900 |
| 1993 | 782 | 252 | 713 | 315 | 3041 | 164 | 278 | 285 | 71 | 100 | 2 | 2 | 26004 |
| 1994 | 668 | 561 | 702 | 516 | 1084 | 147 | 581 | 225 | 49 | 79 | 0 | 1 | 14613 |
| 1995 | 854 | 980 | 514 | 362 | 627 | 115 | 253 | 392 | 116 | 194 | 1 | 0 | 4408 |
| 1996 | 640 | 633 | 1160 | 341 | 536 | 47 | 104 | 534 | 68 | 96 | 0 | 10 | - 4171 |
| 1997 | 438 | 391 | 529 | 311 | 906 | 5 | 112 | 340 | 41 | 36 | 0 | 55 | 3164 |
| 1998 | 428 | 365 | 466 | 325 | 714 | 88 | 151 | 153 | 32 | 9 | 0 | 13 | 3743 |
| 1999 | 387 | 148 | 275 | 256 | 1747 | 133 | 226 | 62 | 26 | 16 | 1 | 31 | 3308 |
| 2000 | 409 | 170 | 463 | 459 | 1767 | 54 | 198 | 76 | 52 | 7 | 0 | 38 | 3693 |
| 2001 | 733 | 178 | 377 | 283 | 1744 | 71 | 256 | 63 | 50 | 6 | 1 | 154 | 3916 |
| 2002 | 452 | 101 | 367 | 186 | 1184 | 141 | 323 | 106 | 183 | 0 | 0 | 277 | - 3322 |

Table 2.3.2.2. The North-east arctic COD stock's consumption of various prey species in 1984-2002 ( 1000 tonnes),
based on Russian consumption calculations.

| Year | Other | Amphipods |  | Shrimp C | Capelin | Herring | Polar cod | Cod | Haddock | Redfish | G. halibut B | Blue whiting | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 536 | - 14 | 44 | - 277 | 546 | 22 | 8 | 13 | 45 | 130 | 0 | 4 | 1639 |
| 1985 | 701 | 238 | 18 | - 172 | 922 | 22 | 0 | 103 | 25 | 69 | 0 | 17 | 2287 |
| 1986 | 602 | 289 | 40 | 114 | 760 | 39 | 88 | 32 | 109 | 115 | 1 | 5 | 2393 |
| 1987 | 539 | 295 | 44 | 179 | 160 | 7 | 67 | 33 | 2 | 95 | 0 | 12 | 1433 |
| 1988 | 585 | 99 | 137 | 100 | 251 | 14 | 0 | 16 | 100 | 96 | 0 | 0 | 1397 |
| 1989 | 518 | - 188 | 118 | 84 | 663 | 3 | 21 | 21 | 2 | 117 | 0 | 0 | 1735 |
| 1990 | 412 | -17 | 53 | 194 | 1150 | 52 | 5 | 20 | 15 | 162 | 0 | 36 | 2117 |
| 1991 | 370 | - 52 | 33 | 210 | 3475 | 30 | 33 | 53 | 23 | 112 | 4 | 7 | 4401 |
| 1992 | 940 | - 19 | 146 | - 206 | 1698 | 428 | 89 | 66 | 42 | 100 | 1 | 0 | 3734 |
| 1993 | 807 | 100 | 83 | 162 | 2496 | 190 | 104 | 139 | 165 | 32 | 6 | 4 | 4288 |
| 1994 | 604 | 145 | 291 | 309 | 1265 | 96 | 247 | 305 | 74 | 47 | 0 | 2 | 3384 |
| 1995 | 875 | 271 | 301 | 371 | 611 | 212 | 111 | 436 | 132 | 98 | 3 | 0 | 3421 |
| 1996 | 656 | - 235 | 734 | 163 | 499 | 99 | 53 | 447 | 71 | 66 | 0 | 7 | 3030 |
| 1997 | 515 | 85 | 386 | - 207 | 527 | 56 | 83 | 409 | 33 | 37 | 3 | 3 | 2343 |
| 1998 | 493 | 115 | 379 | 206 | 657 | 67 | 80 | 148 | 23 | 18 | 0 | 25 | 2211 |
| 1999 | 275 | 43 | 263 | 192 | 1264 | 64 | 82 | 56 | 13 | 13 | 1 | 26 | 2291 |
| 2000 | 334 | 69 | 248 | 269 | 1437 | 46 | 85 | 60 | 24 | 4 | 0 | 22 | 2600 |
| 2001 | 486 | - 47 | 246 | 246 | 1393 | 85 | 89 | 60 | 46 | 3 | 3 | 120 | 2822 |
| 2002 | 356 | - 12 | 233 | 157 | 1687 | 39 | 167 | 114 | 146 | 4 | 0 | 122 | 3037 |

Consumption per cod by cod age group (kg/year), based on Norwegian consumption calculations.

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Table
2.3.2.3
Year/Age

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1997
1998
1999
2000
2001
2002
Consumption per cod by cod age group (kg/year), based on Russian consumption calculations.




~ $\quad$ 아 N No







Table 2.3.4.1. Consumption by minke whale and harp seal (thousand tonnes). The figures for minke whales are based on data from 1992-1995, while the figures for harp seals are based on data for 1990-1996.

| Prey | Minke whale consumption | Harp seal consumption <br> (low capelin stock) | Harp seal consumption <br> (high capelin stock) |
| :--- | :---: | :---: | :---: |
| Capelin | 142 | 23 | 812 |
| Herring | 633 | 394 | 213 |
| Cod | 256 | 298 | 101 |
| Haddock | 128 | 47 | 1 |
| Krill | 602 | 550 | 605 |
| Amphipods | 0 | 304 | $313^{2}$ |
| Shrimp | 0 | 1 | 1 |
| Polar cod | 1 | 880 | 608 |
| Other fish | 55 | 622 | 406 |
| Other crustaceans | 0 | 356 | 312 |
| Total | 1817 | 3491 | 3371 |

${ }^{1}$ the prey species is included in the relevant 'other' group for this predator.
${ }^{2}$ only Parathemisto


Figure 2.1.1.1. Main surface currents of the Nordic and Barents Seas.


Figure 2.1.1.2 Temperature anomalies (upper panel) and salinity anomalies (lower panel) in the section Fugløya Bear Island (Asplin and Dahl, 2003).


Figure 2.1.1.3. Standard Sections and fixed oceanographic stations surveyed by the Institute of Marine Research, Bergen. The University of Bergen is responsible for station M, while the Kola Section is operated by PINRO, Murmansk (Anon. 2001).


Figure 2.2.1. Average zooplankton biomass $\left(\mathrm{g} \mathrm{m}^{-2}\right)$ together with biomass of one-year-old and older capelin (million tonnes) during 1984 - 2002, in the Barents Sea (from Dalpadado et al. 2002, updated with data for 2001-2002).


Figure 2.4.1.1. Winter (December-March) North Atlantic Oscillation index (NAO).


Figure 2.4.1.2. Temperature $\left({ }^{\circ} \mathrm{C}\right)$ and salinity observed during July/August, in the core of Atlantic Water beyond the shelf edge in the sections Svinøy - NW, Gimsøy - NW and Sørkapp - W, averaged between 50 and 200 m depth and horizontally over three stations across the core.


Figure 2.4.1.3. Temperature and salinity in the sections Svinøy - NW and Gimsøy - NW, observed during March/April, in the core of Atlantic Water near the shelf edge, averaged between 50 and 200 m depth and horizontally over three stations across the core.


Figure 2.4.1.4. Time-series of area (blue, in $\mathrm{km}^{2}$ ) and averaged temperature (red) of Atlantic water in the Svinøy section, observed in March/April (upper figure) and July/August (lower figure) 1978-2002.


Figure 2.4.2.1. Distribution of chlorophyll $a$ at $10-\mathrm{m}$ depth during the year at Weather Station Mike in 2002.


Figure 2.4.2.2. Year-to-year variation in the different phases of the development of phytoplankton at Weather Station Mike in the period 1991 to 2002. Diamonds: winter phase; squares: pre-bloom phase; triangles: spring bloom. Continuous lines represent the average for each period. Broken lines represent one standard deviation for each period.


Figure 2.4.3.1 Zooplankton biomass (dry weight) in the upper 200 m in May. A: Arctic influenced water (salinity $<35$, west of $1.4^{\circ} \mathrm{E}$ ). B: Atlantic water (salinity $>35$ ). B: Norwegian Coastal water (salinity $<35$, west of $1.4^{\circ} \mathrm{E}$. Error bars: $95 \%$ confidence limits.


Figure 2.4.3.2 Zooplankton biomass in July-August in the eastern Norwegian Sea ( $0-200 \mathrm{~m}$ ). Integrated biomass within a fixed geographical region divided by its area.


Figure 2.4.4.1 Individual weight-to-length ratio (herring condition index) for Norwegian spring-spawning herring. Data from November and December for herring $30-35 \mathrm{~cm}$ body length. Error bars: $95 \%$ confidence limits.


Figure 2.4.4.2 Zooplankton biomass (dry weight) in Atlantic water in the Norwegian Sea in May (0-200 m) and herring condition index (individual weight-to-length ratio, November and December, 30-35 cm). Error bars: $95 \%$ confidence limits. Linear regression: Condition $=0.0045 *$ biomass $+0.7605 . R^{2}=0.3434$.


Figure 2.4.5.1 Zooplankton biomass in July (year n) vs. zooplankton biomass in May (year n+1).


Figure 2.4.5.2 Average North Atlantic oscillation index (NAO) during four two-month periods (year n) vs. zooplankton biomass in May (year n+1). Circle: prediction of zooplankton biomass in May 2003 based on equation (1).


Figure 2.4.5.3 Herring condition index (year $\mathrm{n}+1$ ) vs. average NAO during four two-month periods (year n). Circle: prediction of herring condition in 2003 based on equation (2).


Figure 2.5.1.1 Temperature (upper panel) and salinity (middle panel) deviations on the Siglunes section off the central north coast of Iceland 1952-2000. The bottom panel shows the variation of zooplankton biomass.


Figure 2.6.1 Horizontal temperature distribution, ${ }^{\circ} \mathrm{C}$, at $10-\mathrm{m}$ depth.


Figure 2.6.2 Horizontal temperature distribution, ${ }^{\circ} \mathrm{C}$, at $400-\mathrm{m}$ depth.


Figure 2.6.3 Vertical distribution of temperature $\left({ }^{\circ} \mathrm{C}\right)$ and salinity in a section at the shelf edge at the Porcupine Bank at $53^{\circ} 30^{\prime} \mathrm{N}$. Station numbers at the top of the panels.


Figure 2.6.4 Vertical distribution of temperature $\left({ }^{\circ} \mathrm{C}\right)$ and salinity in a section from the Faroes to Shetland (NolsøFlugga). Station numbers at the top of the panels.


Figure 2.6.5 Yearly mean temperature and salinity from $50-600 \mathrm{~m}$ (crosses) of all stations in a box with bottom depth $>600 \mathrm{~m}$, west of the Porcupine bank bounded by $52^{\circ}$ to $54^{\circ} \mathrm{N}$ and $16^{\circ}$ to $14^{\circ} \mathrm{W}$. Dotted lines are drawn at plus-minus one standard deviation of all observations in each box, each year.

## $3.1 \quad$ General

### 3.1.1 Stock definition

The Norwegian spring-spawning herring is a unit stock that is characterized by extended migrations, a high number of vertebrae, by large size, special scale characteristics and year-class distribution. The morphological characteristics of this stock resembled those of Icelandic spring-spawning herring. However, the latter stock has disappeared.

ICES areas IIA, IIB and I constitute the distribution area. The adult individuals of the Norwegian spring-spawning herring have a distinct annual migration pattern in the Norwegian Sea. This migration pattern changes over time; at present the herring winters in fjord areas in Northern Norway, spawns on the Norwegian coast (mainly between $62^{\circ}$ and $71^{\circ} \mathrm{N}$ ), and feeds in the Norwegian Sea. The immature stock is distributed mainly in the Barents Sea, but some herring have their nursery area on the Norwegian coast. Historically, for instance in the period 1900-1950, the main spawning areas were located on the Norwegian coast south of $60^{\circ} \mathrm{N}$ (ICES area IV), but at present only a small amount of herring migrate to these spawning grounds. The adult component seldom mixes with herring of other stocks. However, this can occur with immature herring (in certain fjord areas and in the eastern Barents Sea). In this case genetic characteristics are used as a supplement to the above-mentioned separation criteria.

### 3.1.2 ACFM advice and management applicable to 2002 and 2003

In 2001 ACFM stated that "the stock is harvested slightly above $\mathbf{F}_{\mathrm{pa}}=0.15$. The stock biomass is within safe biological limits. The recruitment of the very strong 1992 year class led to an increase in SSB in 1997 to 9 million t , but this has since declined to approximately 6.0 million t in 2001. Continued fishing under the present management agreement, and given the recruitment prospects, gives a low probability of the spawning stock falling below $\mathbf{B}_{\mathrm{pa}}(5.0$ million t$)$ in the medium term. ICES advises that this fishery should be managed according to the agreed management plan, corresponding to a catch of 853000 t in 2002".

In 2002 ACFM stated that "the stock is inside safe biological limits. The stock is harvested at or slightly below $\mathbf{F}_{\mathrm{pa}}=$ 0.15. The recruitment of the very strong 1992 year class led to an increase in SSB in 1997 to approximately 9 million t , but SSB has since declined to just over 5 million $t$ in 2001. The incoming year classes 1998 and 1999 are estimated to be strong. ICES advises that this fishery should be managed according to the agreed management plan, corresponding to a catch of 710000 t in 2003".

At the meeting on Fisheries Consultation on the management of Norwegian spring-spawning herring (AtlantoScandian) herring stock in Harstad, Norway in October 2001, the coastal states (European Union, Faroe Islands, Iceland, Norway, and Russia) agreed to limit their catches to 850000 t in 2002, and there was also agreement on the allocation of the TAC. The agreed TAC was in agreement with ACFM advice.

At the corresponding annual meeting in St Petersburg, Russia in October 2002 the Parties did not come to agreement on a final TAC and the allocation of this. However, most parties agreed to set preliminary national quotas based on a TAC of 710000 t for 2003 and a preliminary allocation based on the 2002 agreement. This preliminary TAC was in agreement with ACFM advice. The basis for the ACFM advice is the agreed management plan (Section 3.7). Catch in numbers for the years 1950-2002 are given in Table 3.2.1.6.

### 3.1.3 Fishery

The catches of Norwegian spring-spawning herring by all countries in 2002 by ICES rectangles are shown in Figure 3.1.2.1 (total whole year) and in Figure 3.1.2.2 (per quarter). In 2002 the catch provided as catch by rectangle represented approximately 763969 tonnes or $94.8 \%$ of the total catch. In general the development of the international fishery shown by these figures follows the known migration pattern for Norwegian spring-spawning herring. The migration pattern, together with environmental factors, was mapped in 2002 during the ICES PGSPFN (Working Group on Surveys on Pelagic Fish in the Norwegian Sea) investigations (ICES 2002/D:07 Ref ACFM, ACME).

Denmark: The Danish fishery of Norwegian spring-spawning herring is carried out mostly by purse seiners and most of the landings were landed in Norway as in previous years. The fishery started in May and ended in the beginning of July and the fishery was carried out in the Norwegian Sea and in the Jan Mayen area (app. 19000 t). In 2002 a fishery was carried out in the area just north of the borderline between the North Sea and Division IIa. The landings from this
fishery have probably consisted of North Sea herring and are not included in the Norwegian Spring-spawning herring figures.

The Faroes: The Faroese herring fishery ( 8 vessels) started in late February 2002 in the Norwegian EEZ (IIa), relatively close to the coast off Sunnmøre (Norwegian coast, $62-63^{\circ} \mathrm{N}$ ) and continued in that area in early March, with some catches farther off the coast in mid-March. In mid-May the fishery was resumed in international waters from 69$73^{\circ} \mathrm{N}$ (ICES Division IIa), and continued in the international area and in the Jan Mayen zone ( $71-74^{\circ} \mathrm{N}$ ) in early June, and the summer fishery terminated in late June in the Svalbard area $\left(75-76^{\circ} \mathrm{N}\right.$, ICES Division IIb). The autumn fishery started in the southern part of the Svalbard zone and in the north-eastern part of the international zone close to the Norwegian zone in late August (ICES Division IIb and IIa), and later in August the fishery had moved into the Norwegian zone off Vesterålen ( $69-70^{\circ} 30^{\prime} \mathrm{N}$ ). The fishery continued in this area until mid-October. All catches were taken with purse-seine. The total catches taken by the Faroese fleet was 32302 t in 2002.

France: France reported no catches in 2002.

Germany: The information from the German fishery was restricted to the amount and location of catches.

Iceland: The Icelandic fishery in 2002 began in the third week of May both in the Jan Mayen zone and in international waters mostly from $69^{\circ} \mathrm{N}$ to $71^{\circ} \mathrm{N}$, between $05^{\circ} \mathrm{E}$ and $02^{\circ} \mathrm{W}$. The next few weeks the fishery moved, first further west into the Jan Mayen zone and then northeast through the international zone and by the end of June had completely moved into the Spitzbergen area. All through July the fishery took place in the Spitzbergen area from $76^{\circ} 20^{\prime} \mathrm{N}$ to $77^{\circ} 30^{\prime} \mathrm{N}$ and between $7^{\circ} \mathrm{E}$ and $12^{\circ} \mathrm{E}$. In the first half of August the fleet started to move south following the herring migration and was fishing in the southern part of the Spitzbergen zone i.e. between $74^{\circ} \mathrm{N}$ and $75^{\circ} \mathrm{N}$ and from $08^{\circ} \mathrm{E}$ to $11^{\circ} \mathrm{E}$ by the end of the month. In the first half of September the fishery gradually noved into the Norwegian EEZ northwest of Lofoten where the fleet fished until the end of the season at the end of September, apart from a few small catches taken in the international zone in the last week of September. A total of 31 fishing vessels took part in the fishery. The total Icelandic catch of Norwegian spring-spawning herring was 127197 tonnes of which only about 47\% were taken with purse seine and $53 \%$ with pelagic trawl. The importance of the pelagic trawl in the fishery is rapidly increasing as only about $25 \%$ of the total catch in 2001 was taken in pelagic trawl. The bulk of the catch (about $68 \%$ ) was taken in June in the Jan Mayen zone and international waters and about $13 \%$ was taken in May in the same area. In each month of July and August 9-10 000 tonnes ( $7-8 \%$ of the catch) were fished, and about 5500 tonnes in September.

Ireland: Ireland reported no catches in 2002.

Netherlands: Catches of Norwegian spring-spawning herring by the Netherlands are taken by pelagic freezer trawlers in the $2^{\text {nd }}$ quarter of the year. The main catches in 2002 originate from area II and are landed frozen for human consumption. The share of the international TAC for the Netherlands in 2002 was 9210 t .

Norway: The Norwegian fishery is carried out by many size categories of vessels. Of the total national quota of $484500 \mathrm{t}, 51 \%$ is allocated to purse seiners, $9 \%$ to trawlers and $39 \%$ to smaller coastal purse seiners. By far the larger part of the Norwegian fishery takes place in northern Norwegian coastal waters (Vestfjorden area) where the herring winters from September until March. Here the herring occurs in concentrations that are easily available to the fishery. In 2002 approximately 140000 t were caught in the wintering area in Northern Norway in January-February, and 48000 t in the spawning area on the Norwegian coast in February-March. Of this 3000 tonnes were caught on the traditional spawning areas south of $60^{\circ} \mathrm{N}$ (ICES area IV). Only 3000 t were caught in the spring/summer fishery in the Norwegian Sea, and the remaining part of the Norwegian quota (approximately 295000 t ) were taken in the period SeptemberDecember on the herring migrating to, and wintering in, the wintering areas in Northern Norway. Approximately $90 \%$ of the Norwegian catches were utilized for human consumption.

Poland: The information on the Polish fishery was restricted to official catch.

Russia: In 2002 the Russian fishery started within the shelf region of the Norwegian EEZ, near Trena Bank (approximately $66^{\circ} \mathrm{N}$ ) at the beginning of February and Sclinna Bank (approximately $65^{\circ} \mathrm{N}$ ) and Budgrunnen Bank (approximately $63^{\circ} \mathrm{N}$ ) at the end of this month. In March the fishing was in progress in the same regions. In February and March the catch was 17520 t . In May-June the commercial vessels conducted fishing in the northern part of the international area in the Norwegian Sea. In May-June the catch was 4865 t . In July-August vessels caught herring in the international area in the Norwegian Sea in the Polar Front region and the zone of Spitsbergen. In September Russian vessels followed the southward migrating fish and continued their fishery in the Norwegian EEZ. In September the fishery of the herring was prolonged in the EEZ of Norway. The herring migrated southwestwards, along the depths of the continental slope. In July-September the catch was 87582 t . In November Russian fishery finished in the

Norwegian EEZ north of the Lofoten area where 3800 t were caught. The Russian fishery is carried out by many types of vessels. The entire Russian catch was utilized for human consumption.

Sweden: The information on the Swedish fishery was restricted to official catch.

UK (Scotland): The information from the Scottish fishery was restricted to the amount and location of catches.

### 3.1.4 Ecosystem aspects

Juveniles and adults of this stock form an important part of the ecosystem in the Barents Sea and the Norwegian Sea. The herring has an important role as transformer of the plankton production to higher trophic levels (cod, seabirds and marine mammals).

### 3.2 Data

### 3.2.1 Commercial catch

The total annual catches of Norwegian spring-spawning herring for the period 1972-2002 (2002 preliminary) are presented in Tables 3.2.1.1 (by fishery) and 3.2.1.2 (by country).

The Working Group noted that in this fishery an unaccounted mortality caused by fishing operations and underreporting probably exists. In general, it was not possible to assess the magnitude of these extra removals from the stock, and taking into account the large catches taken in recent years, the relative importance of such additional mortality is probably low. Therefore, no extra amount to account for these factors has been added in 1994 and later years. In previous years, when the stock and the quotas were much smaller, an estimated amount of fish was added to the catches (Table 3.2.1.1).

In Røttingen et al (2002) the possibility of unaccounted mortality concerning the water content in herring deliveries and errors concerning conversion factors between fillets and live weight of herring were discussed. The paper gave an assessment of underreporting in 2001 of live weight of 18000 tonnes, which is $3 \%$ of the total catch of Norwegian spring-spawning herring landed in Denmark and Norway in 2001 (588 000 tonnes). This underreporting element is probably valid also for other pelagic stocks. For 2003 Norway has increased the allowed subtraction for water content from the weighted catches to $13 \%$. However, for 2004 EU and Norway have agreed to set this subtraction to $2 \%$.

Due to the changing framework of these administrative routines, and to uncertainty on the appropriate period the landing figures should be modified, it was decided not to add any amount to the official landing figures as presented in this year's report. However, this question should be reviewed at the next Working Group meeting.

The combination of national catch-at-age and weight-at-age data for 2002 to obtain the total international catch-at-age and weight-at-age was done using the computer programme SALLOC, a standard ICES software. The official catch, sampled catch, and catch as used by the Working Group, together with number of samples, catch-at-age, and weight-atage for each fishery are given in Tables 3.2.1.3 and 3.2.1.4.

The Working Group noted that not all nations participating in the international fishery for Norwegian spring-spawning herring in 2002 had carried out an adequate sampling of their fishery. The allocation of catches for which no samples were taken and the final catch-at-age and weight-at-age by ICES area is given in Table 3.2.1.5. In general one used the Norwegian age distribution and weights for un-sampled fisheries in the Norwegian Sea in quarter 1-2, and the Russian age distributions and weight keys for quarter 3-4 for un-sampled fisheries in quarter 3-4. The Russian age distribution in quarter 3-4 was calculated using Russian length samples and the Norwegian age-length key for quarter 3 and 4.

In the years 1994-2001 the size group information from each Norwegian catch of herring reported to the Norwegian fishermen's sales organisation for pelagic fish was used to calculate the Norwegian catch in number. In 2002 these data were not available for the WG; instead data from biological sampling from commercial catch were used.

### 3.2.2 Biological data

The natural mortalities and weights in stock (Table 3.2.2.1) and catches (Table 3.2.2.2) used in the assessments are given in Sections 3.3 and 3.4.

### 3.2.3 Surveys

### 3.2.3.1 Spawning grounds

There was no acoustic survey to estimate the abundance of herring in the spawning areas in 2002. Earlier estimates are listed in Table 3.2.3.1.

### 3.2.3.2 Wintering areas

The wintering area of the herring (Vestfjord, Ofotfjord, Tysfjord) was surveyed acoustically by the R/V "Johan Hjort" in November-December 2002. The abundance estimate obtained is given in Table 3.2.3.2. The herring was only observed in the Ofotfjord and Tysfjord in 2002 as opposed to the prior years when herring also was observed in large parts of the Vestfjord. Due to changes in the migration pattern of the younger fraction of the herring stock, with large portions of the 1998 and 1999 year classes wintering off Vesterålen and northwards towards the Tromsøflaket, the estimate of these year classes are strong underestimates.

### 3.2.3.3 Feeding areas

The feeding areas in the Norwegian Sea were surveyed acoustically by the ICES coordinated herring survey (PGSPFN) during the period 25th of April to 29th of May 2002 (ICES 2002/D:07). A complete set of the PGSPFN reports from the years 1995-2002 is found on www.imr.no/PGSPFN. The abundance estimate is given in Table 3.2.3.3.

### 3.2.3.4 Nursery area

The nursery areas of the Norwegian spring-spawning herring are Norwegian fjords and coastal areas, and in the Barents Sea. Since 1988, when the 1983 year class spawned for the first time, the latter area has increased in importance as a nursery area for the herring.

Results from the Russian acoustic survey in the Barents Sea in May-June 2002 are given in Table 3.2.3.4.1. This year the Working Group decided to include data on immature herring obtained during the Russian bottom survey in OctoberDecember 2002 in estimating the younger year classes (A. Krysov, WD). The results from this survey are given in Table 3.2.3.4.2. The results from the 0 -group herring survey in Norwegian Fjords and Coastal areas are given in Table 3.2.3.4.3 and the results from the joint Norwegian-Russian 0-group survey in the Barents Sea are given in Table 3.2.3.4.4.

### 3.2.3.5 Herring larval survey

The larval survey in 2003 was carried out during the period 13-21 April. The survey started in the northern area of the distribution (at approximately $70^{\circ} \mathrm{N}$ ) and proceeded southwards. The distribution of the herring larvae is given in Figure 3.2.3.5. This year, as was the case in 2002, a high proportion of the larvae were found in the northern part of the investigated area, the only area with high densities of herring larvae was found at Røstbanken $\left(68^{\circ} \mathrm{N}\right)$. South of $62^{\circ} \mathrm{N}$ few larvae were recorded, indicating only a minor spawning in that area. Most of the larvae were in the first post yolksack stages and the mean length was 14.7 mm , the highest recorded since the start of these investigations in 1985. The mean age of the larvae was approx. 3 weeks.

The total number of larvae was estimated to $3.7^{*} 10^{12}$, a large reduction compared to the previous years (Table 3.2.3.5). The reason for this reduction is not known. The Working Group has regarded the larvae index as an indicator of the amount of spawning products, and included it as a tuning series. However, no other data and information on the spawning stock and distribution indicates that the spawning in 2003 was reduced to such a degree compared with later years.

### 3.2.4 Other relevant data

### 3.2.4.1 Tagging data

With the exception of 1999 and 2001, tagging has been carried out annually since 1975. The tagging experiments in 2002 were carried out in November in the wintering area (Tysfjorden, $68^{\circ} 05.5^{\prime} \mathrm{N}, 16^{\circ} 22.3^{\prime} \mathrm{E}$ ) where a total of 31993 herring were tagged. During the tagging process, the length of each tagged herring is measured. For each purse seine
catch that is used for tagging, a sample of 100 fish is taken to determine the age distribution within each length group. The age composition of tagged herring in this batch is then estimated from the age distribution in the sample.

Recovery of tags from supervised detector plants has continued, as well as recovery from the standard magnets in the production line of fish processing plants and from individuals. For stock assessment purposes, tags are only used from supervised detector plants where detector efficiency has been tested, and where it is known that the detectors have been working as intended. Two factories filled these criteria in 2002, and a total of 47596 million herring were screened in these factories. Magnet efficiency was $100 \%$ in 2002. All tagged herring recovered were sent to the Norwegian Research Institute where they were measured, weighed and aged. In 2002, 84 tags from herring that were four years or more when tagged, were recovered from factories where detector efficiencies had been working as intended (Table 3.2.4.1).

In 2002 , one of the catches screened had been taken in the southern spawning area at Karmøy (approximately $59^{\circ} \mathrm{N}$ ). The size of the catch was 452 tonnes and 15 tagged herring were found in this catch. This is a large proportion of the total recovered tags (18\%). Of these 15 herrings, 14 had been released in the southern spawning area, indicating nonrandom mixing of tags in the stock that migrated to the southern spawning grounds.

### 3.2.4.2 Prognosis of herring condition

A prognosis of the herring condition index for 2004 was used to estimate weight-at-age for the herring in 2004 in the short-term prognosis for NSSH in SeaStar. The prognosis of the herring condition was derived from a regression of herring condition during the time period 1990 to present on a two-month mean of the NAO index with a one-year time lag as described in Section 2.4.5. Figure 3.2.4.3 shows the relationship between observed and modelled data of the herring condition.

### 3.3 Assessment models: SeaStar

The changes in SeaStar and the use of SeaStar for assessing the Norwegian spring-spawning herring stock is summarised in the text table below:

|  | Analysis of young herring not <br> present in the main tuning series | Tuned year classes |
| :--- | :--- | :--- |
| 2002 | Regressions of VPA to Barents Sea <br> surveys and 0-group data | $1983,1990,1991,1992$ and 1993 |
| 2003 | Tuning in a separate step | $1983,1990,1991,1992,1993$ and |

A documentation of the model was presented (Tjelmeland, WD). The model is written in Mathematica and the documentation text as well as the code can be viewed on the site www.assessment.imr.no. An important change of the model with respect to the two previous years is that the assessment of the young fish that are not present in the main tuning series is not performed by establishing regressions between the final VPA and survey and 0-group indices in the Barents Sea, but performed as a separate tuning step using data (0-group index and acoustic surveys) from the Barents Sea only. In this step, however, all year classes must be present in the tuning since there is no catch data that may reflect in any way the strength of the cohort.

SeaStar is especially designed for Norwegian spring-spawning herring and has been used for the assessment of this stock the latest years. It incorporates a wide variety of information about the stock. Traditionally, using tag data has been important for the management of this stock, and SeaStar incorporates tag data. SeaStar is statistically founded (the objective function expresses a likelihood for the observations), so that information from different sources can be merged without any subjective weighting problem. However, SeaStar does not get information about stock size from signals in the catch data, for which reason the WG this year chose to present both SeaStar and ISVPA as two equally valid models for assessing the stock.

One distinguishing feature of SeaStar is that one may choose to represent only the largest year classes with free parameters (terminal F-values) in the tuning. The rationale for this is that for this stock which has especially strong recruitment dynamics, large relative uncertainty in the weak year classes can generate relatively large uncertainty in the strong year classes. This effect can be serious when the abundance of the small year classes is determined by one or two scales only in certain surveys and certain years of catch data. Figure 3.3.1 shows the results from a simulation
experiment (SeaStar documentation, WD by S . Tjelmeland) that illustrates how uncertainty is propagated from weak to strong year classes. In the experiment, 3 strong and 3 weak year classes were generated repeatedly and an assessment was made tuning on only the large year classes and tuning on all year classes. The figure shows that the uncertainty in the sum of the large year classes is appreciably larger when all year classes are tuned together than when the large year classes are tuned separately, giving support for using the SeaStar approach for assessing fish stocks with strong recruitment dynamics.

The youngest year class in the main tuning last year was the 1993 year class, and at the present meeting the WG included also the intermediate strong year class 1996.

The catchabilities of the acoustic surveys are assumed to have no age dependence or abundance dependence. The reason for this choice is that the mechanism for the recruiting of herring to the surveys is that of migration, not of growth, where the latter process could be assumed to be more regular and modellable. Instead, for each survey there is set a minimum age for inclusion of each year class.

The input data used are:

## Catch data

Acoustic surveys
Tag data
Larval data
0 -group data

Table 3.2.1.5
Table 3.2.3.1, 3.2.3.2, 3.2.3.3, 3.2.3.4
Table 3.2.4.1
Table 3.2.3.5
Table 3.2.3.4.4

The larva index for 2003 is very low (Table 3.2.3.5). The WG decided to leave it out of the analysis. The tag return data for 2002 were also suspicious with 15 tags from one single catch of about 400 tonnes taken at the extreme south of the spawning migration, and consequently the WG decided to leave out also the last year of tag return data.

The acoustic surveys used are (the numbering is used elsewhere in the text of this Section):

1 The spawning grounds along the Norwegian coast
2 The wintering area in Vestfjorden in November/December
3 The wintering area in Vestfjorden in January
4 The young herring in the Barents Sea in May

5 The feeding areas in the Norwegian Sea in May
6 The joint IMR-PINRO capelin survey in September

Minimum age: 5
Minimum age: 4
Minimum age: 5
Minimum age: 1
Maximum age: 2
Minimum age: 3
Minimum age 1
(except the 2002 year class, see below).

For the last of these surveys an extra point for the 2002 year class as 0 -year-old was added this year. This observation stems from a demersal survey in October/November conducted annually by PINRO (Krysov, WD). In order to include this single point in the analysis, the observation was included into the September series with appropriate scaling due to natural mortality during two months. It should be noted, however, that this procedure is not strictly appropriate since it may violate catchability and distributional assumptions. The acoustic survey in Vestfjorden in 2002 found far less 1998 year class than expected from earlier observations, but a considerable amount of this year class was found outside of the survey area immediately after the survey (Jens Chr. Holst, pers. comm.). This data point is therefore omitted from the analysis.

The extremely high estimate of the 1999 year class as 1-year-old fish in the Russian May survey in the Barents Sea was not omitted, however, as the WG had no exogenous information.

It is assumed that the distribution of the main tuning series of older fish follows a gamma distribution with a common CV, which is estimated and that the distribution of the acoustic data in the Barents Sea follows a lognormal distribution. From plots of quantiles of CDF values (not shown) it appears that the former assumption is slightly better than assuming a log-normal distribution, and the latter assumption is much better than assuming a gamma distribution. The tag return data are assumed to follow a Poisson distribution, which is commonly used for rare events, the larval data are
assumed to follow a gamma distribution with an estimated CV , and the zero group data are assumed to follow a normal distribution with an estimated standard deviation.

It has been experienced that when the fish get old the scales get difficult to read and more scales get discarded. This introduces a bias in the age distribution. An attempt has been made to correct for this age bias (Schweder and Tjelmeland, WD).

Previously it has been observed in this WG that as the 1983 and 1985 year classes grew older than about 13 years the age readers tended to transfer fish from the 1983 to the 1985 year class. This problem is part of the general age reading problem for older fish, but has been corrected for separately.

### 3.3.1 Exploratory runs

As has been the case during previous assessments of Norwegian spring-spawning herring with SeaStar, a number of exploratory runs were initially performed to see the effect of various options and settings. These runs were:

## Label Explanation

Default With respect to the 2002 assessment the 1996 year class was included in the tuning.
The other runs were deviations from the run Default:
NoTagsNoLarvae The tag and larvae data were left out to make the model as comparable to ISVPA as possible.
CorrectAgeBias The age bias has been corrected
Youngest1998 The 1998 year class was included in the tuning.
EstimateMortality The natural mortality both of adult herring and young herring in the Barents Sea was estimated.

Lognormal The lognormal distribution was assumed for the main tuning series.
OnlySurvey1
The acoustic series other than the survey on the spawning grounds were omitted.
OnlySurvey2 The acoustic series other than the survey in November/December in Vestfjorden were omitted.

OnlySurvey3 The acoustic series other than the survey in January in Vestfjorden were omitted.
OnlySurvey5 The acoustic series other than the survey on the feeding area in the Norwegian Sea were omitted.

The estimated parameters, spawning stock biomass in 2003 and 2002, and the contribution to the likelihood function from the various data sources are shown in Table 3.3.1.1 for the exploratory runs, excluding those using only one survey, and Table 3.3.1.2 for the runs where only one acoustic survey was used.

The variation in the spawning stock, between 4.6 and 7.8 million tonnes is rather small, giving confidence in the stability of the model. Using only the acoustic survey in the feeding areas from the main survey indices on adult herring while tag data, larval data and the Barents Sea surveys on young herring were retained, gave a 1 million tonnes decrease in the spawning stock, while using only one of the other surveys gave a decrease between 0.3 and 1 million tonnes. There is only a slight difference between using the gamma distribution and lognormal distribution for these surveys. When the 1998 year class is in the tuned year classes, the spawning stock decreases about 1 million tonnes. However, a series of exploratory runs around this option showed that the decrease when using only the survey in the feeding areas and corresponding increase when using only the other surveys was much larger, giving the impression that the year classes should have been present in the surveys for some years in order to give a reliable time trend signal.

When a correction of the age reading bias was attempted, the spawning stock was increased by 0.5 million tonnes, showing the potential seriousness of this source of error. However, the WG felt that the method applied was somewhat premature for making this run the final run from SeaStar. The higher uncertainties in the Barents Sea surveys (4 and 6) are reflected in their smaller contribution to the log-likelihood per data point.

The natural mortality for older herring was estimated at 0.13 by the model, as opposed to 0.14 last year. However, the estimated mortality is close to the mortality of 0.15 assumed by the WG. The natural mortality for the young herring in the Barents Sea was estimated at 0.5 , which is lower than the mortality of 0.9 assumed by the WG. A bootstrap was not
done for this run, but the latter estimate is probably much more uncertain than the former, since there is no mortality signal from the tag data involved.

The log-likelihood per data point are at the same level for all the main tuning surveys (surveys $1,2,3$ and 5), while the log-likelihood per data point from the surveys in the Barents Sea is much lower, reflecting the larger uncertainty in these surveys.

The catchability of the survey in the feeding areas (cat5) is estimated at about 1 ( 1.03 for the main run last year) in all runs, reflecting that this survey is the survey that is in closest agreement with the model's perception of stock abundance.

### 3.3.2 Final assessment including uncertainty analysis and retrospective analysis

Figure 3.3.2.1 shows the fit of the VPA scaled with the estimated catchability to the observations in the survey on the spawning grounds for the tuned year classes. Figure 3.3.2.2 shows a similar fit for the December survey in Vestfjorden, Figure 3.3.2.3 for the January survey in Vestfjorden and Figure 3.3.2.4 for the feeding areas in the Norwegian Sea. With exception of the 1991 year class in the December survey and possibly the 1983 year class in the survey on the spawning grounds all year classes seem to recruit well to all surveys. The fit is reasonably good for all surveys. However, the 1996 year class seems to be somewhat systematically underrepresented in the survey on the feeding areas. It is interesting to note that in 1999 all year classes increased with respect to the previous year in the January survey and the survey on the spawning grounds, as all year classes also did in the September survey in 1998.

Figure 3.3.2.5 shows the retrospective plot. The effect of the anomalous survey observations to the assessment in 1999 is clearly reflected in the retrospective run. The general trend is well followed, but there is a larger spread in the recent years than was the case for the retrospective plot last year. The reason for this is probably the 3-year span between the youngest tuned year class entered this year and the next youngest tuned year class. There is a tendency for consistently overestimating the stock.

Figure 3.3.2.6 shows a histogram of the spawning stock in 2003 for the bootstrap replicates. The standard deviation is about 1 million tonnes - about the same as last year, giving a CV of 0.17 . However, the distribution is skewed, with the maximum likelihood value being about 0.3 million tonnes larger than the mean and about 0.45 million tonnes larger than the median. Table 3.3.2 shows a summary of the parameters, spawning stock and total spawning stock from the bootstrap replicates. Figure 3.3.2.7 shows a scatter plot where the covariation between pairs of parameters, between parameters and spawning stock, and between parameters and total stock have been plotted. The parameters are labelled with shortened names, given in Table 3.3.2.

### 3.3.3 Recruitment estimation

The VPA from the Default run of SeaStar made back to 1950 (Table 3.3.3) gives a series of the spawning stock and recruitment as 0 -year-old herring that may be used to build a recruitment function to drive medium-term and long-term simulations. Figure 3.3 .3 shows a scatter plot of spawning stock and recruitment labelled with the recruitment years. $25 \%$ of the most successful recruitment points (marked in red) has been omitted from a fit of the Beverton-Holt recruitment relation (line) and are treated separately during simulations. There seems to be a large stochastic variation. However, there is a definite signal from the spawning stock.

### 3.3.4 Short-Term Projection

Table 3.3.4.1 shows the input data to the short-term projection. For the weight-at-age in 2004 the weight-at-age in 2003, which is taken from December 2002 data, was multiplied with the ratio of the condition factor in December 2003 to the condition factor in December 2002. The condition factors are described in Section 2 and in Section 3.2. As exploitation pattern was used a least-square fit of a third-order polynomial to the F-values generated by SeaStar in 2002. It is assumed that a catch of 0.710 million tonnes is taken in 2003. The weight-at-age in the catch is taken as the mean of the weight-at-age in the catch in the period 2000-2002. The mature fraction-at-age in 2002 was used throughout the shortterm period. The number-at-age in 2003 is taken from the SeaStar Default run. The recruitment at age 0 in 2003-2005 is set to zero and consequently the total stock biomass is an underestimate.

Table 3.3.4.2 shows the results of the short-term projection. For a population-weighted fishing mortality over the ages $5-14$ of about the agreed value of 0.125 the expected catch in 2004 is about 0.850 million tonnes. The spawning stock will increase to 2004 because of the 1999 year class being nearly fully mature. For the agreed fishing mortality the
spawning stock will decrease to about 5.6 million tonnes in 2005, a development that is in accordance with projections made by this WG last year.

The population-weighted fishing mortality for ages 5-14, Fbar, is in 20030.11 , which is generated by an assumed catch of 0.71 million tonnes. If the exploitation pattern used by ACFM last year is used in the projection, Fbar in 2003 is calculated to 0.12 and the catch corresponding to an Fbar of 0.125 in 2004 is about 0.76 million tonnes.

### 3.4 Assessment models: ISVPA

### 3.4.1 Exploratory runs

For NSS herring stock assessment the catch-controlled version of the ISVPA with constraint of unbiased model approximation of logarithmic catch-at-age was used. Current program realization of the ISVPA, which allows simultaneous implementation of up to 7 age-structured indexes (of mature, immature or total stock) and up to 3 integrated indexes of SSB, was reported to the ICES WG on methods (ICES 2003/ D:03).

The catch-controlled version of ISVPA is considered to be more appropriate for stocks with very variable selection pattern, like NSS herring and is therefore used here. It is, however, less resistant to noise in the data. Therefore the median of the distributions of squared residuals (MDN) in the ISVPA loss function was used to increase robustness.

Profiles of the components of the ISVPA loss function, coming from each survey and from separable description by the model of catch-at-age data, are shown in Figures 3.4.1.and 3.4.2. In the catch-controlled version of the model, for which the abundances are calculated directly via catches, residuals in catch-at-age could be considered as a measure of separability of estimated fishing mortalities, but not as a measure of deviations of estimated abundances from "real" ones. For all runs (except specially mentioned) age groups 1-16+ were used in analysis.

As it can be seen from Figure 3.4.1, when SSE was used as a measure of goodness of fit for the model to age-structured surveys, only survey N 2 (acoustic survey in wintering areas) revealed a good minimum, while it was somewhat weaker for survey N4 (acoustic survey in feeding area). The application of MDN instead of SSE (Fig. 3.4.2) resulted in the appearance of distinct minima for all indexes, except index N4.

Since index N4 is considered to be one of the most valuable observations and gives the best fit to abundance estimates from the cohort part of the model, but gives an uncertain signal about the value of terminal effort factor (very flat minimum), it was decided to pay more attention to this index. Assuming that a low quality of signal from a good source of information might be the result of a hindering effect in the model of the signal from the age-structured survey data, the abundance-at-age $I_{a, y}$ values of this survey were " detrended" by the following procedure:

$$
I_{a, y}(\text { new })=\frac{I_{a, y}}{\sum_{a=1}^{a=o l d e s t} I_{a, y}}
$$

Resulting ("detrended") data now contain information only about relative (normalized to sum=1 for each year) structure of stock abundance.

In order not to lose information about trend in the stock biomass, which was present in the initial data, the data on abundance-at-age of this survey were recalculated to the SSB index as:

$$
I_{y}^{S S B}=\sum_{a=1}^{a=o l d e s t}\left(I_{a, y} w_{a, y}^{\text {instock }} \text { maturity }_{a, y}\right)
$$

This data set containing SSB indices, derived from survey N4, was used as auxiliary information in tuning of the model.
It turned out that data from survey N4, split into a "detrended" relative index of stock age structure, and an index of SSB, revealed distinct minima (see Fig. 3.4.2) and approximately at the same place with respect to the terminal effort factor $f_{\text {term }}$. Now all sources of information gave almost coherent signals.

Figure 3.4.3 represents profiles of the ISVPA total loss function for different cases with respect to treatment of survey N4 data. As it can be seen, global minimum is not strongly changed by the choice of how to take the information from this survey into account, but difference between solutions in terms of SSB (Fig. 3.4.4) between some of the cases exists.

The properties of the proposed procedure of "detrending" of age-structured survey data was tested using a simulated "noisy" data set, generated at the ICES Working Group on Methods of Stock Assessment (ICES 2003/ D:03). For analysis the index with trend in catchability was used. ISVPA was implemented in the same version, as for herring -catch-controlled, unbiased approximation of logarithmic catch-at-age, minimization of MDN. In the ISVPA loss function only the component for this survey was used. "Detrending" of the index data gave the result, apparently closer to the "truth" in comparison to tune on "initial" data (see Figures 3.4.5 and 3.4.6).

Since the procedure of "splitting" the survey information is not traditional for the WG and needs further consideration in relation to similar approaches used in some other models (i.e. BORMICON), it was decided to use the index N4 with SSE-minimization, which produces a somewhat better pronounced minimum in comparison to MDN.

Experiments with changing the range of ages, included into the analysis, were also made and presented by Sigurd Tjelmeland. Exclusion of older ages or (and) age group 1 from analysis in some cases changed the result substantially. For example, exclusion of age group 1 with simultaneous restricting older ages from $16+$ to $15+$ (with the same other settings - survey N4 is used as "detrended" + trend in SSB ) brought down the estimate of SSB(2002) from 6 million to 4.7 million tonnes, while almost not changing the pattern of residuals in logarithmic catches (Fig. 3.4.7).

It was proposed (Sigurd Tjelmeland) to include "dual" selection pattern into the model (such potential possibility is reserved in the ISVPA, but for two successive continuous periods). In contrast, for herring it was proposed to use two specific selection patterns for specific years - according to whether a strong or an ordinary generation is entering the fishery at age 4 . Only preliminary runs were done, but in some of them the estimated values of abundance for young age groups became closer to the results of the Sea Star model. This approach may be fruitful in merging positive features of the Sea Star and ISVPA models.

### 3.4.2 Final assessment, including uncertainty analysis and retrospective analysis

The ISVPA options, taken in the basic run in comparison to those previously used are listed in the table below.

| Settings/options for the ISVPA run | $\mathbf{2 0 0 2}(\mathbf{W G )}$ | $\mathbf{2 0 0 2}$ (Coastal <br> States meeting) | $\mathbf{2 0 0 3}$ (WG) |
| :--- | :--- | :--- | :--- |
| Numbers of age structured tuning series | 0 | 6 | 6 |
| Version of the model | Catch-controlled | Catch-controlled | Catch-controlled |
| Numbers of biomass tuning series | 0 | 0 | 0 |
| Constraint | Unbiased model <br> description of <br> logarithmic <br> catch-at-age | Unbiased model <br> description of <br> logarithmic <br> catch-at-age | Unbiased model <br> description of <br> logarithmic <br> catch-at-age |
| Number of years with a separable constraint | $1986-2001$ | 1986-2001 | $1986-2002$ |
| Reference age for separable constraint | Yo | No | no |
| Constant selection pattern | Equal to that for <br> previous age | Equal to that for <br> previous age | Equal to that for <br> previous age |
| S on the last age | $5-14$ | $5-14$ | $5-14$ |
| Age span for calculation of reference F | 1 | 1 | 1 |
| Weight given to age groups and years in separable <br> period | NA | Constant, q(a) <br> are estimated | Constant, q(a) <br> are estimated |
| Catchability model for all fleets | $2-16+$ | $1-16+$ |  |
| Age range for the analysis | NA | SSE | Yes <br> Survey weights for all ages in all fleets <br> What is minimized for residuals in logarithmic catch-at- <br> age |
| NA | SSE for survey <br> on feeding area <br> (N4) |  |  |

What is minimized for logarithmic abundance-at-age (for indexes)

MDN; absolute
median deviation (AMD)

| MDN; absolute <br> median deviation <br> (AMD) | MDN |
| :--- | :--- |

A summary of the results of the stock assessment is given in Table 3.4.1. Estimates of abundance and instantaneous fishing mortality coefficients by years and ages are given in Tables 3.4.2-3.4.3. Since age group 1 is absent in the catch-at-age matrix in 2002 and age group 2 appears in it for the first time the values for these age groups, estimated by the model, are to be substituted in the short-term forecast by 538000 for age group 1 and 3935000 for age group 2, taken directly from results of young fish surveys in the Barents Sea.

The retrospective analysis (see Fig. 3.4.8), undertaken with the same options independently of whether signals from all surveys in restricted year ranges are still existing or not, showed low stability of results, except for runs with 2000 and 2001 terminal years, for which the estimates were very close to each other.

The bootstrap procedure for the catch-controlled version of ISVPA was somewhat different in comparison with the effort-controlled one. For the catch-controlled version, for which abundance is calculated directly using catches and estimated selection at age is used only for terminal ages and in terminal year, approximation of catch-at-age data by the model cannot be considered as measure of errors in the data. That is why instead of resampling of residuals (as it is done for the effort-controlled version), new catch-at-age data were generated simply by adding a lognormally distributed noise with variance 0.3 . For auxiliary information the same procedure was used, variance for each survey was also taken equal to 0.3 .

Results in terms of SSB are apparently shifted downwards in final years (Fig. 3.4.9). This may be explained by complete deterioration of the solution for some combinations of noise in the data (disappearance of minima of components of the ISVPA loss function), which was revealed in approximately $10 \%$ of the runs and was expressed in extremely high estimates of "best" values of the effort factor in the terminal year ( $f_{\text {term }}>7$ ), compared to $f_{\text {term }}=2.57$ in the basic run. Perhaps the procedure of noising of input data needs further consideration for the catch-controlled version of the model. Better knowledge of real values of variances for survey data is also needed.

### 3.4.3 Recruitment estimation

No specific procedure for recruitment estimation was used. For the short-term projection recruits were estimated as the geometrical mean of estimates for 1986-2001, taken from the basic run of the ISVPA.

### 3.4.4 Short-Term Projection

Input data for the short-term projection are given in Table 3.4.4 and the results are shown in Table 3.4.5.

### 3.5 Comparison between Seastar and ISVPA

### 3.5.1 Evaluation and comparison of the models ISVPA and SeaStar

The Working Group discussed in detail and made many trial runs with the ISVPA and SeaStar models. This discussion is reflected in Sections 3.3 and 3.4. A summary of some inherent properties is given in the text table below:

| Element | ISVPA | SeaStar |
| :--- | :--- | :--- |
| Changes in model compared <br> with last years estimate | Includes tuning with survey data | New method of estimating year <br> classes not included in the tuning. In <br> the 2002 assessment the 1993 year <br> class was the youngest year class in <br> the tuning, in 2003 the 1996 year <br> class was the youngest |
| Information on tags and <br> larvae | Not included | Included |
| Information from fisheries | Separability is a central element, one <br> constant exploitation pattern | No information from the catch signal |


| Data for estimating <br> recruitment | Catch data with tuning to acoustic <br> surveys of immature herring | Acoustic surveys on immature <br> herring, data from 0-group trawl <br> indexes |
| :--- | :--- | :--- |
| Estimation method | Minimization of the median of log <br> residuals | Maximum likelihood |

Since 2001 model work has been undertaken to investigate different properties and the models have been modified to increase transparency and to make them more equal. This work should continue.

The Working Group concluded that both models are relevant and applicable to assess the state of the Norwegian springspawning herring. The main difference of the models is the estimation of the year classes entering into the spawning stock. This is a period with dynamic changes, the herring migrates from the nursery area in the Barents Sea to the Norwegian Sea, the maturing process begins and the fish "migrates" into the fishery (fishing in the Barents Sea is not possible due to area closure in the Russian EEZ and minimum length of 25 cm in the Norwegian EEZ). The amount of herring migrating to the northern Norwegian Sea varies from year to year according to year-class strength. This year-toyear variation may be important to take into account when discussing the separability issue. SeaStar lacks information from this phase, while ISVPA relies on constant selection pattern.

### 3.5.2 Comparison of ISVPA and SeaStar assessments

The results from the two models are not in agreement with regard to the younger year classes. This was also the case in 2002. Fig 3.5.2.1 gives the values for the different year classes in the assessment in 2002 compared with the perception of the same year classes in the 2003 assessment.

In the 2003 assessment the main discrepancy is the size of the 1998 year class. The Working Group evaluated data on the abundance of the year classes 1992 and 1998 from various sources (acoustic surveys, trawl indexes, VPA estimates). The relevant data on these year classes are given in Table 3.5.2. The Working Group concluded that the 1998 year class as 4 -year-olds is not at the same level of abundance as the 1992 year class as 4 -year-olds. The input data on year-class strength and selection used for the short-term projection to the two models are shown in Fig. 3.5.2.2 and Fig. 3.5.2.3.

### 3.6 Biological reference points

The process of establishing biological reference points (including a $\mathbf{B}_{\text {lim }}$ value of 2.5 million tonnes) for this stock is given in Røttingen (2000). The SGPRP03 stated that the segmented regression was significant at a $5 \%$ level and gave a change point of 2.3 million tonnes. Furthermore, the SGPRP03 has the opinion that these numbers are close to each other and as this stock is managed to an agreed harvest control rule, there is no need to change the $\mathbf{B}_{\mathrm{lim}}$ value.

If the spawning stock falls below 5.0 million tonnes $\left(\mathbf{B}_{\mathrm{pa}}\right)$ the managers have, according to the harvest control rule, decided on a linear reduction in fishing mortality from the agreed maximum fishing mortality 0.125 ( $\mathbf{F}_{\mathrm{pa}}$ is set to 0.15 ) at $\mathbf{B}_{\mathrm{pa}}$ to 0.05 at $\mathbf{B}_{\mathrm{lim}}$. The WGNPBW agrees on the conclusion of the SGPRP03 on the $\mathbf{B}_{\mathrm{lim}}$, and is, due to the elements in the agreed harvest control rule, also of the opinion that there is no reason at present to change the values for the biological reference points $\mathbf{F}_{\mathrm{pa}}$ and $\mathbf{B}_{\mathrm{pa}}$.

### 3.7 Management targets

EU, Faroe Islands, Iceland, Norway, and Russia agreed to implement a long-term management plan. This plan consists of the following elements (ICES 2002/CRR:255):

1. Every effort shall be made to maintain a level of Spawning Stock Biomass (SSB) greater than the critical level ( $\boldsymbol{B}_{\text {lim }}$ ) of $2500000 t$.
2. For the year 2001 and subsequent years, the Parties agreed to restrict their fishing on the basis of a TAC consistent with a fishing mortality rate of less than 0.125 for appropriate age groups as defined by ICES, unless future scientific advice requires modification of this fishing mortality rate.
3. Should the SSB fall below a reference point of $5000000 t\left(\boldsymbol{B}_{p a}\right)$, the fishing mortality rate, referred under paragraph 2, shall be adapted in the light of scientific estimates of the conditions to ensure a safe and rapid
recovery of the SSB to a level in excess of 5000000 t. The basis for such an adaptation should be at least a linear reduction in the fishing mortality rate from 0.125 at $\boldsymbol{B}_{p a}\left(5000000\right.$ t) to $0.05 \boldsymbol{B}_{\text {lim }}(2500000 t$ ).
4. The Parties shall, as appropriate, review and revise these management measures and strategies on the basis of any new advice provided by ICES.

ICES considers that the objectives of this agreement are consistent with the precautionary approach.

Table 3.2.1.1 Catches of Norwegian spring-spawning herring (tonnes) since 1972.

| Year | A | $B^{1}$ | C | D | Total | Total catch used in WG |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | - | 9895 | 3,266 ${ }^{2}$ | - | 13,161 | 13,161 |
| 1973 | 139 | 6,602 | 276 | - | 7,017 | 7,017 |
| 1974 | 906 | 6,093 | 620 | - | 7,619 | 7,619 |
| 1975 | 53 | 3,372 | 288 | - | 3,713 | 13,713 |
| 1976 | - | 247 | 189 | - | 436 | 10,436 |
| 1977 | 374 | 11,834 | 498 | - | 12,706 | 22,706 |
| 1978 | 484 | 9,151 | 189 | - | 9,824 | 19,824 |
| 1979 | 691 | 1,866 | 307 | - | 2,864 | 12,864 |
| 1980 | 878 | 7,634 | 65 | - | 8,577 | 18,577 |
| 1981 | 844 | 7,814 | 78 | - | 8,736 | 13,736 |
| 1982 | 983 | 10,447 | 225 | - | 11,655 | 16,655 |
| 1983 | 3,857 | 13,290 | 907 | - | 18,054 | 23,054 |
| 1984 | 18,730 | 29,463 | 339 | - | 48,532 | 53,532 |
| 1985 | 29,363 | 37,187 | 197 | 4,300 | 71,047 | 169,872 |
| 1986 | $71,122^{3}$ | 55,507 | 156 | - | 126,785 | 225,256 |
| 1987 | 62,910 | 49,798 | 181 | - | 112,899 | 127,306 |
| 1988 | 78,592 | 46,582 | 127 | - | 125,301 | 135,301 |
| 1989 | 52,003 | 41,770 | 57 | - | 93,830 | 103,830 |
| 1990 | 48,633 | 29,770 | 8 | - | 78,411 | 86,411 |
| 1991 | 48,353 | 31,280 | 50 | - | 79,683 | 84,683 |
| 1992 | 43,688 | 55,737 | 23 | - | 99,448 | 104,448 |
| 1993 | 117,195 | 110,212 | 50 | - | 227,457 | 232,457 |
| 1994 | 288,581 | 190,643 | 4 | - | 479,228 | 479,228 |
| 1995 | 320,731 | 581,495 | 0 | - | 902,226 | 902,226 |
| 1996 | 462,248 | 758,035 | 0 | - | 1,220,283 | 1,220,283 |
| $1997{ }^{5}$ |  |  | 0 | - | 1,426,507 | 1,426,507 |
| $1998{ }^{5}$ |  |  | 0 | - | 1,223,131 | 1,223,131 |
| $1999{ }^{6}$ |  |  | 0 | - | 1,235,433 | 1,235,433 |
| $2000{ }^{7}$ |  |  | 0 | - | 1,207,201 | 1,207,201 |
| $2001{ }^{8}$ |  |  | 0 | - | 770,066 | 770,066 |
| $2002{ }^{9}$ |  |  | 0 | - | 806,086 | 806,086 |

$\mathrm{A}=$ catches of adult herring in winter
$B=$ mixed herring fishery in remaining part of the year
$\mathrm{C}=$ by-catches of 0 - and 1-group herring in the sprat fishery
$\mathrm{D}=$ USSR-Norway by-catch in the capelin fishery (2-group)

[^0]Table 3.2.1.2 Total catch of Norwegian spring-spawning herring (tonnes) since 1972. Data provided by Working Group members.

| Year | Norway | $\begin{aligned} & \text { USSR/ } \\ & \text { Russia } \end{aligned}$ | Denmark | Faroes | Iceland | Ireland | Netherlands | Greenland | UK | Germany | France | Poland | Sweden | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 13,161 | - | - | - | - | - | - | - | - | - | - | - | - | 13,161 |
| 1973 | 7,017 | - | - | - | - | - | - | - | - | - | - | - | - | 7,017 |
| 1974 | 7,619 | - | - | - | - | - | - | - | - | - | - | - | - | 7,619 |
| 1975 | 13,713 | - | - | - | - | - | - | - | - | - | - | - | - | 13,713 |
| 1976 | 10,436 | - | - | - | - | - | - | - | - | - | - | - | - | 10,436 |
| 1977 | 22,706 | - | - | - | - | - | - | - | - | - | - | - | - | 22,706 |
| 1978 | 19,824 | - | - | - | - | - | - | - | - | - | - | - | - | 19,824 |
| 1979 | 12,864 | - | - | - | - | - | - | - | - | - | - | - | - | 12,864 |
| 1980 | 18,577 | - | - | - | - | - | - | - | - | - | - | - | - | 18,577 |
| 1981 | 13,736 | - | - | - | - | - | - | - | - | - | - | - | - | 13,736 |
| 1982 | 16,655 | - | - | - | - | - | - | - | - | - | - | - | - | 16,655 |
| 1983 | 23,054 | - | - | - | - | - | - | - | - | - | - | - | - | 23,054 |
| 1984 | 53,532 | - | - | - | - | - | - | - | - | - | - | - | - | 53,532 |
| 1985 | 167,272 | 2,600 | - | - | - | - | - | - | - | - | - | - | - | 169,872 |
| 1986 | 199,256 | 26,000 | - | - | - | - | - | - | - | - | - | - | - | 225,256 |
| 1987 | 108,417 | 18,889 | - | - | - | - | - | - | - | - | - | - | - | 127,306 |
| 1988 | 115,076 | 20,225 | - | - | - | - | - | - | - | - | - | - | - | 135,301 |
| 1989 | 88,707 | 15,123 | - | - | - | - | - | - | - | - | - | - | - | 103,830 |
| 1990 | 74,604 | 11,807 | - | - | - | - | - | - | - | - | - | - | - | 86,411 |
| 1991 | 73,683 | 11,000 | - | - | - | - | - | - | - | - | - | - | - | 84,683 |
| 1992 | 91,111 | 13,337 | - | - | - | - | - | - | - | - | - | - | - | 104,448 |
| 1993 | 199,771 | 32,645 | - | - | - | - | - | - | - | - | - | - |  | 232,457 |
| 1994 | 380,771 | 74,400 | - | 2,911 | 21,146 | - | - | - | - | - | - | - | - | 479,228 |
| 1995 | 529,838 | 101,987 | 30,577 | 57,084 | 174,109 | - | 7,969 | 2,500 | 881 | 556 | - | - | - | 905,501 |
| 1996 | 699,161 | 119,290 | 60,681 | 52,788 | 164,957 | 19,541 | 19,664 | - | 46,131 | 11,978 | - | - | 22,424 | 1,220,283 |
| 1997 | 860,963 | 168,900 | 44,292 | 59,987 | 220,154 | 11,179 | 8,694 |  | 25,149 | 6,190 | 1,500 |  | 19,499 | 1,426,507 |
| 1998 | 743,925 | 124,049 | 35,519 | 68,136 | 197,789 | 2,437 | 12,827 |  | 15,9711 | 7,003 | 605 | - | 14,863 | 1,223,131 |
| 1999 | 740,640 | 157,328 | 37,010 | 55,527 | 203,381 | 2,412 | 5,871 | - | 9,207 | - | - |  | 14,057 | 1,235,433 |
| 2000 | 713,500 | 163,261 | 34,968 | 68,625 | 186,035 | 8,939 | - | - | 14,096 | 3,298 | - | - | 14,749 | 1,207,201 |
| 2001 | 495,036 | 109,054 | 24,038 | 34,170 | 77,693 | - | 6,439 |  | 12,230 | 1,588 | - | - | 9,818 | 770,066 |
| $2002^{1}$ | 487,233 | 113,763 | 18,998 | 32,302 | 127,197 | - | 9,392 | - | 3,482 | 3,017 | - | 1,226 | 9,486 | 806,086 |

${ }^{1}$ Preliminary, as provided by Working Group members.

Table 3.2.1.4. Weight (kg) at age by country.


Table 3.2.1.5

Summary of Sampling by Country

AREA : IVa

| Country | Sampled <br> Catch | Official <br> Catch | No. of <br> samples | No. <br> measured | No. <br> aged |
| :---: | :---: | :---: | :---: | :---: | :---: |

AREA : IIb

| Country | Sampled <br> Catch | Official <br> Catch |
| :--- | ---: | ---: |
| Sweden | 0.00 | 280.00 |
| Russia | 5044.00 | 5044.00 |
| Iceland | 60726.00 | 64892.00 |
| Germany | 0.00 | 1005.00 |
| Faroes | 0.00 | 3451.00 |
| Total IIb | 65770.00 | 74672.00 |
|  |  | 74672.00 |
| Sum of Offical Catches |  | 0.00 |

AREA : IIa

| Country | Sampled Catch | $\begin{gathered} \text { Official } \\ \text { Catch } \end{gathered}$ | No. of samples | No. measured | No. aged | SOP |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK (Scot) | 0.00 | 3482.00 | 0 | 0 | 0 | 0.00 |
| Sweden | 0.00 | 9206.00 | 0 | 0 | 0 | 0.00 |
| Russia | 104919.00 | 108719.00 | 61 | 17763 | 300 | 75.97 |
| Poland | 0.00 | 1226.00 | 0 | 0 | 0 | 0.00 |
| Norway | 463823.00 | 485118.00 | 52 | 4698 | 3710 | 100.00 |
| Netherlands | 9392.00 | 9392.00 | 21 | 1737 | 525 | 88.04 |
| Iceland | 51333.00 | 62305.00 | 3 | 112 | 112 | 0.00 |
| Germany | 1303.00 | 2012.00 | 33 | 12118 | 429 | 111.29 |
| Faroes | 0.00 | 28851.00 | 0 | 0 | 0 | 0.00 |
| Denmark | 12491.00 | 18998.00 | 4 | 249 | 249 | 114.68 |
| Total IIa | 643261.00 | 729309.00 | 174 | 36677 | 5325 | 88.23 |
| Sum of Offical Catches : |  | 729309.00 |  |  |  |  |
| Unallocat Working | : | $\begin{array}{r} 0.00 \\ 729309.00 \end{array}$ |  |  |  |  |

Table 3.2.1.5 continued

PERIOD : 1

| Country | Sampled Catch | Official Catch |
| :---: | :---: | :---: |
| UK (Scot) | 0.00 | 3482.00 |
| Russia | 17520.00 | 17520.00 |
| Norway | 186489.00 | 188451.00 |
| Germany | 1303.00 | 1303.00 |
| Faroes | 0.00 | 7211.00 |
| Denmark | 0.00 | 6507.00 |
| Period Total | 205312.00 | 224474.00 |
| Sum of Offical Catches :Unallocated Catch : |  | 224474.00 |
|  |  | 0.00 |


| No. of | No. |
| :---: | :---: |
| samples | measured |
| 0 | 0 |
| 6 | 717 |
| 19 | 1764 |
| 33 | 12118 |
| 0 | 0 |
| 0 | 0 |
| 58 | 14599 |


| No. | SOP |
| ---: | ---: |
| aged | $\% \%$ |
| 0 | 0.00 |
| 16 | 0.00 |
| 1494 | 100.00 |
| 429 | 111.29 |
| 0 | 0.00 |
| 0 | 0.00 |
| 1939 | 91.54 |

PERIOD : 2

| Country | Sampled Catch | Official <br> Catch |
| :---: | :---: | :---: |
| Sweden | 0.00 | 9486.00 |
| Russia | 4865.00 | 4865.00 |
| Norway | 2928.00 | 3071.00 |
| Netherlands | 9392.00 | 9392.00 |
| Iceland | 112059.00 | 112059.00 |
| Germany | 0.00 | 1714.00 |
| Faroes | 0.00 | 11886.00 |
| Denmark | 12491.00 | 12491.00 |
| Period Total | 141735.00 | 164964.00 |
| Sum of Offical Catches :Unallocated Catch : |  | 164964.00 |
|  |  | 0.00 |
| Working Group Catch : |  | 164964.00 |


| No. of | No. |
| :---: | :---: |
| samples | measured |
| 0 | 0 |
| 3 | 243 |
| 15 | 1209 |
| 21 | 1737 |
| 4 | 192 |
| 0 | 0 |
| 0 | 0 |
| 4 | 249 |


| No. | SOP |
| ---: | ---: |
| aged | $\%$ |
| 0 | 0.00 |
| 51 | 0.00 |
| 1069 | 100.00 |
| 525 | 88.04 |
| 192 | 0.00 |
| 0 | 0.00 |
| 0 | 0.00 |
| 249 | 114.68 |
| 2086 | 18.01 |

PERIOD : 3

| Country | Sampled Catch | Official Catch |
| :---: | :---: | :---: |
| Russia | 87578.00 | 87578.00 |
| Poland | 0.00 | 1226.00 |
| Norway | 0.00 | 21295.00 |
| Iceland | 0.00 | 15138.00 |
| Faroes | 0.00 | 10967.00 |
| Period Total | 87578.00 | 136204.00 |
| Sum of Offical | hes : | 136204.00 |
| Unallocated Catch |  | 0.00 |

PERIOD : 4

| Country | Sampled <br> Catch | Official <br> Catch |
| :---: | ---: | ---: |
| Russia | 0.00 | 3800.00 |
| Norway | 274406.00 | 274406.00 |
| Faroes | 0.00 | 2238.00 |
| Period Total | 274406.00 | 280444.00 |
|  |  | 280444.00 |
| Sum of Offical Catches : | 0.00 |  |
| Unallocated Catch : |  | 280444.00 |

Table 3.2.1.5 continued
Total over all Areas and Periods

| Country | Sampled Catch | Official Catch | No. of samples | $\begin{aligned} & \text { No. } \\ & \text { measured } \end{aligned}$ | No. aged | $\underset{\%}{\mathrm{SOP}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK (Scot) | 0.00 | 3482.00 | 0 | 0 | 0 | 0.00 |
| Sweden | 0.00 | 9486.00 | 0 | 0 | 0 | 0.00 |
| Russia | 109963.00 | 113763.00 | 70 | 20338 | 300 | 76.63 |
| Poland | 0.00 | 1226.00 | 0 | 0 | 0 | 0.00 |
| Norway | 463823.00 | 487223.00 | 52 | 4698 | 3710 | 100.00 |
| Netherlands | 9392.00 | 9392.00 | 21 | 1737 | 525 | 88.04 |
| Iceland | 112059.00 | 127197.00 | 4 | 192 | 192 | 0.00 |
| Germany | 1303.00 | 3017.00 | 33 | 12118 | 429 | 111.29 |
| Faroes | 0.00 | 32302.00 | 0 | 0 | 0 | 0.00 |
| Denmark | 12491.00 | 18998.00 | 4 | 249 | 249 | 114.68 |
| Total for Stock | 709031.00 | 806086.00 | 184 | 39332 | 5405 | 80.69 |
| Sum of Offical Catches : Unallocated Catch : Working Group Catch : |  | 806086.00 |  |  |  |  |
|  |  | 0.00 |  |  |  |  |
|  |  | 806086.00 |  |  |  |  |

```
DETAILS OF DATA FILLING-IN
```

    Filling-in for record : ( 3)
    Using Only
>> ( 4) Norway
Filling-in for record : ( 5)
Using Only
>> (1) Norway
Filling-in for record : (6)
Using Only
>> ( 2) Norway
Filling-in for record : ( 7 )
Using Only
>> ( 1) Norway
Filling-in for record : ( 8)
Using Only
>> (2) Norway
Filling-in for record : ( 10 )
>> ( 2) Norway
Filing-in for record : ( 10 )
Using Only
>> ( 9) Russia
Filling-in for record : ( 12 )
Using Only
>> (1) Norway
$\gg(1)$ Norway
Filling-in for record : (14)
Using Only
>> (2) Norway
Filling-in for record : ( 15)
Using Only
>> ( 9 ) Russia
Filling-in for record : ( 16 )
$\gg$ ( 9) Russia
Filling-in for record : ( 16 )
Using Only
>> (2) Norway
Filling-in for record : ( 17)
Using Only
>> ( 9) Russia
Filling-in for record : ( 18)
Using Only
>> (2) Norway
Filling-in for record : ( 19)
Using Only
>> (2) Norway
Filling-in for record : ( 20 )
>> (2) Norway
Filling-in for record : ( 20 )
Using Only
>> (1) Norway
Filling-in for record : ( 21 )
Using Only
>> (2) Norway
$\gg(2)$ Norway
Filling-in for record : (22)
Using Only
>> (2) Norway
Filling-in for record : ( 23)
Using Only
$\gg\left(\begin{array}{l}\text { ) Norway }\end{array}\right.$
1 IIa

3 IIa

1 IVa

2 IVa

1 IIa

2 IIa

4 IIa

1 IIa

2 IIa

3 IIa

2 IIb

3 IIb

2 IIa

2 IIb

1 IIa

2 IIa

2 IIb

1 IIa

Table 3.2.1.5 continued
Filling-in for record : ( 24)
Using Only >> ( 9) Russia Filling-in for record : ( 25)
Using Only
>> (1) Norway
Filling-in for record : ( 26)
Using Only >> (2) Norway Filling-in for record : ( 27)
Using Only >> ( 9) Russia Filling-in for record : ( 28)
Using Only
>> ( 9) Russia
Filling-in for record : ( 29)
Using Only
>> ( 2) Norway
Filling-in for record : ( 30 )
Using Only

$$
\gg \quad(9) \text { Russia }
$$

Catch Numbers at Age by Area
-------------------------------

| Ages | IVa | IIb |
| ---: | ---: | ---: |
| 0 | 0.00 | 0.00 |
| 1 | 0.00 | 0.00 |
| 2 | 118.20 | 24354.86 |
| 3 | 251.52 | 68784.79 |
| 4 | 1314.40 | 117105.53 |
| 5 | 708.93 | 28016.71 |
| 6 | 1490.58 | 12859.10 |
| 7 | 34.70 | 4561.69 |
| 8 | 213.61 | 10576.97 |
| 9 | 654.77 | 25649.07 |
| 10 | 1918.32 | 47337.37 |
| 11 | 958.50 | 23643.99 |
| 12 | 157.07 | 1412.86 |
| 13 | 39.25 | 349.51 |
| 14 | 27.08 | 723.95 |
| 15 | 13.54 | 377.94 |

Mean Weight-at-age by Area (Kg)

| Ages | IVa | IIb |  |
| :---: | :---: | :---: | :---: |
| 0 | 0.0000 | 0.0000 |  |
| 1 | 0.0000 |  |  |
| 2 |  | 0.0675 |  |
|  | 0.0000 |  |  |
| 3 | 0.1230 |  |  |
| 4 | 0.1748 | 0.0527 |  |
| 5 | 0.2292 | 0.1173 |  |
| 6 | 0.2648 | 0.2435 |  |
| 7 | 0.2530 | 0.2646 |  |
| 8 | 0.3035 | 0.2921 |  |
| 9 | 0.3058 | 0.2949 |  |
| 10 | 0.3165 | 0.3962 |  |
| 11 | 0.3258 | 0.3239 |  |
| 12 | 0.3594 | 0.3360 |  |
| 13 | 0.4028 | 0.4012 |  |
| 14 | 0.2938 | 0.3238 |  |
| 15 | 0.4240 | 0.4083 |  |

Poland
3 IIa
Faroes
1 IIa
Faroes
2 IIa
Faroes
3 IIa
Faroes
3 IIa
Faroes
2 IIa
Faroes
3 IIa

| IIa | Total |
| ---: | ---: |
| 0.00 | 0.00 |
| 0.00 | 0.00 |
| 37401.58 | 61874.64 |
| 129044.98 | 198081.30 |
| 521902.22 | 640322.19 |
| 225097.64 | 253823.28 |
| 308441.63 | 322791.31 |
| 25173.21 | 29769.59 |
| 82234.61 | 93025.19 |
| 236786.81 | 263090.66 |
| 609012.50 | 658268.19 |
| 312314.06 | 336916.56 |
| 50938.13 | 52508.05 |
| 11946.71 | 12335.47 |
| 6186.32 | 6937.35 |
| 9656.14 | 10047.62 |

IIa |  | Total |
| ---: | ---: |
| 0.0000 | 0.0000 |
| 0.0000 | 0.0000 |
| 0.0587 | 0.0563 |
| 0.1344 | 0.1285 |
| 0.2041 | 0.1968 |
| 0.2568 | 0.2553 |
| 0.2824 | 0.2816 |
| 0.3056 | 0.3034 |
| 0.3263 | 0.3226 |
| 0.3261 | 0.3231 |
| 0.3374 | 0.3350 |
| 0.3468 | 0.3451 |
| 0.3703 | 0.3694 |
| 0.4073 | 0.4071 |
| 0.2880 | 0.2918 |
| 0.4363 | 0.4352 |

> Total

$$
.0000
$$

$$
.0563
$$

$$
.1968
$$

$$
0.2553
$$

$$
\begin{aligned}
& 0.2816 \\
& 0.3034
\end{aligned}
$$

$$
0.3226
$$

$$
\begin{aligned}
& 0.3231 \\
& 0.3350
\end{aligned}
$$

$$
0.3451
$$

$$
0.2918
$$

$$
0.4352
$$

3 IIa

1 IIa

2 IIa

3 IIa

4 IIa

2 IIb

3 IIb





N











茴茳亶

Table 3.2.2.1 Norwegian spring spawning herring. Stock weights at age (in kg )

| year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0.001 | 0.008 | 0.047 | 0.1 | 0.204 | 0.23 | 0.255 | 0.275 | 0.29 | 0.305 | 0.315 | 0.325 | 0.33 | 0.34 |
| 1951 | 0.001 | 0.008 | 0.047 | 0.1 | 0.204 | 0.23 | 0.255 | 0.275 | 0.29 | 0.305 | 0.315 | 0.325 | 0.33 | 0.34 |
| 1952 | 0.001 | 0.008 | 0.047 | 0.1 | 0.204 | 0.23 | 0.255 | 0.275 | 0.29 | 0.305 | 0.315 | 0.325 | 0.33 | 0.34 |
| 1953 | 0.001 | 0.008 | 0.047 | 0.1 | 0.204 | 0.23 | 0.255 | 0.275 | 0.29 | 0.305 | 0.315 | 0.325 | 0.33 | 0.34 |
| 1954 | 0.001 | 0.008 | 0.047 | 0.1 | 0.204 | 0.23 | 0.255 | 0.275 | 0.29 | 0.305 | 0.315 | 0.325 | 0.33 | 0.34 |
| 1955 | 0.001 | 0.008 | 0.047 | 0.1 | 0.195 | 0.213 | 0.26 | 0.275 | 0.29 | 0.305 | 0.315 | 0.325 | 0.33 | 0.34 |
| 1956 | 0.001 | 0.008 | 0.047 | 0.1 | 0.205 | 0.23 | 0.249 | 0.275 | 0.29 | 0.305 | 0.315 | 0.325 | 0.33 | 0.34 |
| 1957 | 0.001 | 0.008 | 0.047 | 0.1 | 0.136 | 0.228 | 0.255 | 0.262 | 0.29 | 0.305 | 0.315 | 0.325 | 0.33 | 0.34 |
| 1958 | 0.001 | 0.008 | 0.047 | 0.1 | 0.204 | 0.242 | 0.292 | 0.295 | 0.293 | 0.305 | 0.315 | 0.33 | 0.34 | 0.345 |
| 1959 | 0.001 | 0.008 | 0.047 | 0.1 | 0.204 | 0.252 | 0.26 | 0.29 | 0.3 | 0.305 | 0.315 | 0.325 | 0.33 | 0.34 |
| 1960 | 0.001 | 0.008 | 0.047 | 0.1 | 0.204 | 0.27 | 0.291 | 0.293 | 0.321 | 0.318 | 0.32 | 0.344 | 0.349 | 0.37 |
| 1961 | 0.001 | 0.008 | 0.047 | 0.1 | 0.232 | 0.25 | 0.292 | 0.302 | 0.304 | 0.323 | 0.322 | 0.321 | 0.344 | 0.357 |
| 1962 | 0.001 | 0.008 | 0.047 | 0.1 | 0.219 | 0.291 | 0.3 | 0.316 | 0.324 | 0.326 | 0.335 | 0.338 | 0.334 | 0.347 |
| 1963 | 0.001 | 0.008 | 0.047 | 0.1 | 0.185 | 0.253 | 0.294 | 0.312 | 0.329 | 0.327 | 0.334 | 0.341 | 0.349 | 0.341 |
| 1964 | 0.001 | 0.008 | 0.047 | 0.1 | 0.194 | 0.213 | 0.264 | 0.317 | 0.363 | 0.353 | 0.349 | 0.354 | 0.357 | 0.359 |
| 1965 | 0.001 | 0.008 | 0.047 | 0.1 | 0.186 | 0.199 | 0.236 | 0.26 | 0.363 | 0.35 | 0.37 | 0.36 | 0.378 | 0.387 |
| 1966 | 0.001 | 0.008 | 0.047 | 0.1 | 0.185 | 0.219 | 0.222 | 0.249 | 0.306 | 0.354 | 0.377 | 0.391 | 0.379 | 0.378 |
| 1967 | 0.001 | 0.008 | 0.047 | 0.1 | 0.18 | 0.228 | 0.269 | 0.27 | 0.294 | 0.324 | 0.42 | 0.43 | 0.366 | 0.368 |
| 1968 | 0.001 | 0.008 | 0.047 | 0.1 | 0.115 | 0.206 | 0.266 | 0.275 | 0.274 | 0.285 | 0.35 | 0.325 | 0.363 | 0.408 |
| 1969 | 0.001 | 0.008 | 0.047 | 0.1 | 0.115 | 0.145 | 0.27 | 0.3 | 0.306 | 0.308 | 0.318 | 0.34 | 0.368 | 0.36 |
| 1970 | 0.001 | 0.008 | 0.047 | 0.1 | 0.209 | 0.272 | 0.23 | 0.295 | 0.317 | 0.323 | 0.325 | 0.329 | 0.38 | 0.37 |
| 1971 | 0.001 | 0.015 | 0.08 | 0.1 | 0.19 | 0.225 | 0.25 | 0.275 | 0.29 | 0.31 | 0.325 | 0.335 | 0.345 | 0.355 |
| 1972 | 0.001 | 0.01 | 0.07 | 0.15 | 0.15 | 0.14 | 0.21 | 0.24 | 0.27 | 0.3 | 0.325 | 0.335 | 0.345 | 0.355 |
| 1973 | 0.001 | 0.01 | 0.085 | 0.17 | 0.259 | 0.342 | 0.384 | 0.409 | 0.404 | 0.461 | 0.52 | 0.534 | 0.5 | 0.5 |
| 1974 | 0.001 | 0.01 | 0.085 | 0.17 | 0.259 | 0.342 | 0.384 | 0.409 | 0.444 | 0.461 | 0.52 | 0.543 | 0.482 | 0.482 |
| 1975 | 0.001 | 0.01 | 0.085 | 0.181 | 0.259 | 0.342 | 0.384 | 0.409 | 0.444 | 0.461 | 0.52 | 0.543 | 0.482 | 0.482 |
| 1976 | 0.001 | 0.01 | 0.085 | 0.181 | 0.259 | 0.342 | 0.384 | 0.409 | 0.444 | 0.461 | 0.52 | 0.543 | 0.482 | 0.482 |
| 1977 | 0.001 | 0.01 | 0.085 | 0.181 | 0.259 | 0.343 | 0.384 | 0.409 | 0.444 | 0.461 | 0.52 | 0.543 | 0.482 | 0.482 |
| 1978 | 0.001 | 0.01 | 0.085 | 0.18 | 0.294 | 0.326 | 0.371 | 0.409 | 0.461 | 0.476 | 0.52 | 0.543 | 0.5 | 0.5 |
| 1979 | 0.001 | 0.01 | 0.085 | 0.178 | 0.232 | 0.359 | 0.385 | 0.42 | 0.444 | 0.505 | 0.52 | 0.551 | 0.5 | 0.5 |
| 1980 | 0.001 | 0.01 | 0.085 | 0.175 | 0.283 | 0.347 | 0.402 | 0.421 | 0.465 | 0.465 | 0.52 | 0.534 | 0.5 | 0.5 |
| 1981 | 0.001 | 0.01 | 0.085 | 0.17 | 0.224 | 0.336 | 0.378 | 0.387 | 0.408 | 0.397 | 0.52 | 0.543 | 0.512 | 0.512 |
| 1982 | 0.001 | 0.01 | 0.085 | 0.17 | 0.204 | 0.303 | 0.355 | 0.383 | 0.395 | 0.413 | 0.453 | 0.468 | 0.506 | 0.506 |
| 1983 | 0.001 | 0.01 | 0.085 | 0.155 | 0.249 | 0.304 | 0.368 | 0.404 | 0.424 | 0.437 | 0.436 | 0.493 | 0.495 | 0.495 |
| 1984 | 0.001 | 0.01 | 0.085 | 0.14 | 0.204 | 0.295 | 0.338 | 0.376 | 0.395 | 0.407 | 0.413 | 0.422 | 0.437 | 0.437 |
| 1985 | 0.001 | 0.01 | 0.023 | 0.148 | 0.234 | 0.265 | 0.312 | 0.346 | 0.37 | 0.395 | 0.397 | 0.428 | 0.428 | 0.428 |
| 1986 | 0.001 | 0.01 | 0.085 | 0.054 | 0.206 | 0.265 | 0.289 | 0.339 | 0.368 | 0.391 | 0.382 | 0.388 | 0.395 | 0.395 |
| 1987 | 0.001 | 0.01 | 0.055 | 0.09 | 0.143 | 0.241 | 0.279 | 0.299 | 0.316 | 0.342 | 0.343 | 0.362 | 0.376 | 0.376 |
| 1988 | 0.001 | 0.015 | 0.05 | 0.098 | 0.135 | 0.197 | 0.277 | 0.315 | 0.339 | 0.343 | 0.359 | 0.365 | 0.376 | 0.376 |
| 1989 | 0.001 | 0.015 | 0.1 | 0.154 | 0.175 | 0.209 | 0.252 | 0.305 | 0.367 | 0.377 | 0.359 | 0.395 | 0.396 | 0.396 |
| 1990 | 0.001 | 0.008 | 0.048 | 0.219 | 0.198 | 0.258 | 0.288 | 0.309 | 0.428 | 0.37 | 0.403 | 0.387 | 0.44 | 0.44 |
| 1991 | 0.001 | 0.011 | 0.037 | 0.147 | 0.21 | 0.244 | 0.3 | 0.324 | 0.336 | 0.343 | 0.382 | 0.366 | 0.425 | 0.425 |
| 1992 | 0.001 | 0.007 | 0.03 | 0.128 | 0.224 | 0.296 | 0.327 | 0.355 | 0.345 | 0.367 | 0.341 | 0.361 | 0.43 | 0.47 |
| 1993 | 0.001 | 0.008 | 0.025 | 0.081 | 0.201 | 0.265 | 0.323 | 0.354 | 0.358 | 0.381 | 0.369 | 0.396 | 0.393 | 0.374 |
| 1994 | 0.001 | 0.01 | 0.025 | 0.075 | 0.151 | 0.254 | 0.318 | 0.371 | 0.347 | 0.412 | 0.382 | 0.407 | 0.41 | 0.41 |
| 1995 | 0.001 | 0.018 | 0.025 | 0.066 | 0.138 | 0.23 | 0.296 | 0.346 | 0.388 | 0.363 | 0.409 | 0.414 | 0.422 | 0.41 |
| 1996 | 0.001 | 0.018 | 0.025 | 0.076 | 0.118 | 0.188 | 0.261 | 0.316 | 0.346 | 0.374 | 0.39 | 0.39 | 0.384 | 0.398 |
| 1997 | 0.001 | 0.018 | 0.025 | 0.096 | 0.118 | 0.174 | 0.229 | 0.286 | 0.323 | 0.37 | 0.378 | 0.386 | 0.36 | 0.393 |
| 1998 | 0.001 | 0.018 | 0.025 | 0.074 | 0.147 | 0.174 | 0.217 | 0.242 | 0.278 | 0.304 | 0.31 | 0.359 | 0.34 | 0.344 |
| 1999 | 0.001 | 0.018 | 0.025 | 0.102 | 0.15 | 0.223 | 0.24 | 0.264 | 0.283 | 0.315 | 0.345 | 0.386 | 0.386 | 0.386 |
| 2000 | 0.001 | 0.018 | 0.025 | 0.102 | 0.15 | 0.223 | 0.24 | 0.264 | 0.283 | 0.315 | 0.345 | 0.386 | 0.386 | 0.386 |
| 2001 | 0.001 | 0.018 | 0.025 | 0.075 | 0.178 | 0.238 | 0.247 | 0.296 | 0.307 | 0.314 | 0.328 | 0.351 | 0.376 | 0.406 |
| 2002 | 0.001 | 0.018 | 0.054 | 0.098 | 0.159 | 0.211 | 0.272 | 0.305 | 0.314 | 0.331 | 0.337 | 0.346 | 0.356 | 0.381 |

Table 3.2.2.2 Norwegian spring spawning herring. Catch weight at age (in kg ).

| year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 16+ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0.007 | 0.025 | 0.058 | 0.11 | 0.188 | 0.211 | 0.234 | 0.253 | 0.266 | 0.28 | 0.294 | 0.303 | 0.312 | 0.32 | 0.323 | 0.331 | 0.335 |
| 1951 | 0.009 | 0.029 | 0.068 | 0.13 | 0.222 | 0.249 | 0.276 | 0.298 | 0.314 | 0.33 | 0.346 | 0.357 | 0.368 | 0.377 | 0.381 | 0.39 | 0.395 |
| 1952 | 0.008 | 0.026 | 0.061 | 0.115 | 0.197 | 0.221 | 0.245 | 0.265 | 0.279 | 0.293 | 0.308 | 0.317 | 0.327 | 0.335 | 0.339 | 0.346 | 0.351 |
| 1953 | 0.008 | 0.027 | 0.063 | 0.12 | 0.205 | 0.23 | 0.255 | 0.275 | 0.29 | 0.305 | 0.32 | 0.33 | 0.34 | 0.347 | 0.351 | 0.359 | 0.364 |
| 1954 | 0.008 | 0.026 | 0.062 | 0.117 | 0.201 | 0.225 | 0.25 | 0.269 | 0.284 | 0.299 | 0.313 | 0.323 | 0.333 | 0.341 | 0.345 | 0.352 | 0.357 |
| 1955 | 0.008 | 0.027 | 0.063 | 0.119 | 0.204 | 0.229 | 0.254 | 0.274 | 0.289 | 0.304 | 0.318 | 0.328 | 0.338 | 0.346 | 0.35 | 0.358 | 0.363 |
| 1956 | 0.008 | 0.028 | 0.066 | 0.126 | 0.215 | 0.241 | 0.268 | 0.289 | 0.304 | 0.32 | 0.336 | 0.346 | 0.357 | 0.365 | 0.369 | 0.378 | 0.383 |
| 1957 | 0.008 | 0.028 | 0.066 | 0.127 | 0.216 | 0.243 | 0.269 | 0.29 | 0.306 | 0.322 | 0.338 | 0.348 | 0.359 | 0.367 | 0.371 | 0.38 | 0.385 |
| 1958 | 0.009 | 0.03 | 0.07 | 0.133 | 0.227 | 0.255 | 0.283 | 0.305 | 0.321 | 0.338 | 0.355 | 0.366 | 0.377 | 0.386 | 0.39 | 0.399 | 0.404 |
| 1959 | 0.009 | 0.03 | 0.071 | 0.135 | 0.231 | 0.259 | 0.287 | 0.31 | 0.327 | 0.344 | 0.36 | 0.372 | 0.383 | 0.392 | 0.397 | 0.406 | 0.411 |
| 1960 | 0.006 | 0.011 | 0.074 | 0.119 | 0.188 | 0.277 | 0.337 | 0.318 | 0.363 | 0.379 | 0.36 | 0.42 | 0.411 | 0.439 | 0.45 | 0.444 | 0.448 |
| 1961 | 0.006 | 0.01 | 0.045 | 0.087 | 0.159 | 0.276 | 0.322 | 0.372 | 0.363 | 0.393 | 0.407 | 0.397 | 0.422 | 0.447 | 0.465 | 0.452 | 0.452 |
| 1962 | 0.009 | 0.023 | 0.055 | 0.085 | 0.148 | 0.288 | 0.333 | 0.36 | 0.352 | 0.35 | 0.374 | 0.384 | 0.374 | 0.394 | 0.399 | 0.411 | 0.416 |
| 1963 | 0.008 | 0.026 | 0.047 | 0.098 | 0.171 | 0.275 | 0.268 | 0.323 | 0.329 | 0.336 | 0.341 | 0.358 | 0.385 | 0.353 | 0.381 | 0.386 | 0.386 |
| 1964 | 0.009 | 0.024 | 0.059 | 0.139 | 0.219 | 0.239 | 0.298 | 0.295 | 0.339 | 0.35 | 0.358 | 0.351 | 0.367 | 0.375 | 0.372 | 0.427 | 0.434 |
| 1965 | 0.009 | 0.016 | 0.048 | 0.089 | 0.217 | 0.234 | 0.262 | 0.331 | 0.36 | 0.367 | 0.386 | 0.395 | 0.393 | 0.404 | 0.401 | 0.429 | 0.437 |
| 1966 | 0.008 | 0.017 | 0.04 | 0.063 | 0.246 | 0.26 | 0.265 | 0.301 | 0.41 | 0.425 | 0.456 | 0.46 | 0.467 | 0.446 | 0.459 | 0.465 | 0.474 |
| 1967 | 0.009 | 0.015 | 0.036 | 0.066 | 0.093 | 0.305 | 0.305 | 0.31 | 0.333 | 0.359 | 0.413 | 0.446 | 0.401 | 0.408 | 0.439 | 0.427 | 0.431 |
| 1968 | 0.01 | 0.027 | 0.049 | 0.075 | 0.108 | 0.158 | 0.375 | 0.383 | 0.364 | 0.382 | 0.441 | 0.41 | 0.442 | 0.517 | 0.491 | 0.464 | 0.487 |
| 1969 | 0.009 | 0.021 | 0.047 | 0.072 | 0.105 | 0.152 | 0.296 | 0.376 | 0.329 | 0.329 | 0.341 | 0.363 | 0.385 | 0.377 | 0.451 | 0.423 | 0.429 |
| 1970 | 0.008 | 0.058 | 0.085 | 0.105 | 0.171 | 0.256 | 0.216 | 0.277 | 0.298 | 0.304 | 0.305 | 0.309 | 0.357 | 0.348 | 0.357 | 0.367 | 0.376 |
| 1971 | 0.011 | 0.053 | 0.121 | 0.177 | 0.216 | 0.25 | 0.277 | 0.305 | 0.333 | 0.353 | 0.366 | 0.377 | 0.388 | 0.399 | 0.419 | 0.444 | 0.444 |
| 1972 | 0.011 | 0.029 | 0.062 | 0.103 | 0.154 | 0.215 | 0.258 | 0.295 | 0.322 | 0.341 | 0.354 | 0.365 | 0.376 | 0.387 | 0.406 | 0.43 | 0.43 |
| 1973 | 0.006 | 0.053 | 0.106 | 0.161 | 0.213 | 0.239 | 0.255 | 0.277 | 0.287 | 0.324 | 0.338 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 |
| 1974 | 0.006 | 0.055 | 0.117 | 0.168 | 0.222 | 0.249 | 0.265 | 0.288 | 0.299 | 0.337 | 0.352 | 0.267 | 0.324 | 0.324 | 0.324 | 0.324 | 0.324 |
| 1975 | 0.009 | 0.079 | 0.169 | 0.241 | 0.318 | 0.358 | 0.381 | 0.413 | 0.429 | 0.484 | 0.506 | 0.384 | 0.466 | 0.466 | 0.466 | 0.466 | 0.466 |
| 1976 | 0.007 | 0.062 | 0.132 | 0.189 | 0.25 | 0.28 | 0.298 | 0.323 | 0.336 | 0.379 | 0.396 | 0.3 | 0.364 | 0.364 | 0.364 | 0.364 | 0.364 |
| 1977 | 0.011 | 0.091 | 0.193 | 0.316 | 0.35 | 0.398 | 0.439 | 0.495 | 0.511 | 0.558 | 0.583 | 0.537 | 0.537 | 0.537 | 0.537 | 0.537 | 0.537 |
| 1978 | 0.012 | 0.1 | 0.21 | 0.274 | 0.424 | 0.454 | 0.495 | 0.524 | 0.596 | 0.613 | 0.65 | 0.59 | 0.59 | 0.59 | 0.59 | 0.59 | 0.59 |
| 1979 | 0.01 | 0.088 | 0.181 | 0.293 | 0.359 | 0.416 | 0.436 | 0.482 | 0.482 | 0.539 | 0.553 | 0.518 | 0.518 | 0.518 | 0.518 | 0.518 | 0.518 |
| 1980 | 0.012 | 0.101 | 0.202 | 0.266 | 0.399 | 0.449 | 0.46 | 0.485 | 0.472 | 0.618 | 0.645 | 0.608 | 0.594 | 0.594 | 0.594 | 0.594 | 0.594 |
| 1981 | 0.01 | 0.082 | 0.163 | 0.196 | 0.291 | 0.341 | 0.368 | 0.38 | 0.397 | 0.436 | 0.45 | 0.492 | 0.481 | 0.481 | 0.481 | 0.481 | 0.481 |
| 1982 | 0.01 | 0.087 | 0.159 | 0.256 | 0.312 | 0.378 | 0.415 | 0.435 | 0.449 | 0.448 | 0.506 | 0.493 | 0.499 | 0.499 | 0.499 | 0.499 | 0.499 |
| 1983 | 0.011 | 0.09 | 0.165 | 0.217 | 0.265 | 0.337 | 0.378 | 0.41 | 0.426 | 0.435 | 0.444 | 0.468 | 0.461 | 0.461 | 0.461 | 0.461 | 0.461 |
| 1984 | 0.009 | 0.047 | 0.145 | 0.218 | 0.262 | 0.325 | 0.346 | 0.381 | 0.4 | 0.413 | 0.405 | 0.426 | 0.415 | 0.415 | 0.415 | 0.415 | 0.415 |
| 1985 | 0.009 | 0.022 | 0.022 | 0.214 | 0.277 | 0.295 | 0.338 | 0.36 | 0.381 | 0.397 | 0.409 | 0.417 | 0.435 | 0.435 | 0.435 | 0.435 | 0.435 |
| 1986 | 0.007 | 0.077 | 0.097 | 0.055 | 0.249 | 0.294 | 0.312 | 0.352 | 0.374 | 0.398 | 0.402 | 0.401 | 0.41 | 0.41 | 0.41 | 0.41 | 0.41 |
| 1987 | 0.01 | 0.075 | 0.091 | 0.124 | 0.173 | 0.253 | 0.232 | 0.312 | 0.328 | 0.349 | 0.353 | 0.37 | 0.385 | 0.385 | 0.385 | 0.385 | 0.385 |
| 1988 | 0.008 | 0.062 | 0.075 | 0.124 | 0.154 | 0.194 | 0.241 | 0.265 | 0.304 | 0.305 | 0.317 | 0.308 | 0.334 | 0.334 | 0.334 | 0.334 | 0.334 |
| 1989 | 0.01 | 0.06 | 0.204 | 0.188 | 0.264 | 0.26 | 0.282 | 0.306 | 0.309 | 0.391 | 0.422 | 0.364 | 0.429 | 0.429 | 0.429 | 0.429 | 0.429 |
| 1990 | 0.007 | 0.078 | 0.102 | 0.23 | 0.239 | 0.266 | 0.305 | 0.308 | 0.376 | 0.407 | 0.412 | 0.424 | 0.428 | 0.428 | 0.428 | 0.428 | 0.428 |
| 1991 | 0.007 | 0.015 | 0.104 | 0.208 | 0.25 | 0.288 | 0.312 | 0.316 | 0.33 | 0.344 | 0.372 | 0.354 | 0.398 | 0.398 | 0.398 | 0.398 | 0.398 |
| 1992 | 0.007 | 0.075 | 0.103 | 0.191 | 0.233 | 0.304 | 0.337 | 0.365 | 0.361 | 0.371 | 0.403 | 0.365 | 0.394 | 0.404 | 0.406 | 0.408 | 0.41 |
| 1993 | 0.007 | 0.03 | 0.106 | 0.153 | 0.243 | 0.282 | 0.32 | 0.33 | 0.365 | 0.373 | 0.379 | 0.38 | 0.385 | 0.39 | 0.395 | 0.4 | 0.405 |
| 1994 | 0.007 | 0.063 | 0.102 | 0.194 | 0.239 | 0.28 | 0.317 | 0.328 | 0.356 | 0.372 | 0.39 | 0.379 | 0.399 | 0.403 | 0.405 | 0.407 | 0.405 |
| 1995 | 0.007 | 0.063 | 0.102 | 0.153 | 0.192 | 0.234 | 0.283 | 0.328 | 0.349 | 0.356 | 0.374 | 0.366 | 0.393 | 0.387 | 0.4 | 0.4 | 0.4 |
| 1996 | 0.007 | 0.063 | 0.136 | 0.136 | 0.168 | 0.206 | 0.262 | 0.309 | 0.337 | 0.366 | 0.36 | 0.361 | 0.367 | 0.379 | 0.379 | 0.379 | 0.379 |
| 1997 | 0.007 | 0.063 | 0.089 | 0.167 | 0.184 | 0.207 | 0.232 | 0.277 | 0.305 | 0.331 | 0.328 | 0.344 | 0.343 | 0.397 | 0.357 | 0.51 | 0.51 |
| 1998 | 0.007 | 0.063 | 0.111 | 0.15 | 0.216 | 0.221 | 0.249 | 0.277 | 0.316 | 0.338 | 0.374 | 0.372 | 0.366 | 0.396 | 0.377 | 0.406 | 0.406 |
| 1999 | 0.007 | 0.063 | 0.096 | 0.173 | 0.228 | 0.262 | 0.274 | 0.292 | 0.307 | 0.335 | 0.362 | 0.371 | 0.399 | 0.396 | 0.4 | 0.4 | 0.404 |
| 2000 | 0.007 | 0.063 | 0.124 | 0.175 | 0.222 | 0.242 | 0.289 | 0.303 | 0.31 | 0.328 | 0.349 | 0.383 | 0.411 | 0.41 | 0.419 | 0.409 | 0.409 |
| 2001 | 0.007 | 0.063 | 0.105 | 0.166 | 0.214 | 0.252 | 0.268 | 0.305 | 0.308 | 0.322 | 0.337 | 0.363 | 0.353 | 0.378 | 0.4 | 0.427 | 0.427 |
| 2002 | 0.0 | 0.063 | 0.056 | 0.128 | 0.198 | 0.255 | 0.281 | 0.303 | 0.322 | 0.323 | 0.334 | 0.345 | 0.369 | 0.407 | 0.41 | 0.435 | 0.435 |

Table 3.2.3.1 Norwegian Spring-spawning herring. Estimates from the acoustic surveys on the spawning stock in February-March. Numbers in millions.

| Year <br> Age | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 |  |  | 101 | 183 | 44 |  |  | 16 |  | 407 |  |  | 106 |
| 3 | 255 | 5 | 187 | 59 |  |  | 128 | 1792 | 231 |  |  | 1516 |  |
| 4 | 146 | 373 | 0 | 54 |  |  | 676 | 7621 | 7638 |  | 381 | 337 | 690 |
| 5 | 6805 | 103 | 345 | 12 |  |  | 1375 | 3807 | 11243 |  | 1905 | 1286 | 164 |
| 6 | 202 | 5402 | 112 | 354 |  |  | 476 | 2151 | 2586 |  | 10640 | 2979 | 592 |
| 7 |  | 182 | 4489 | 122 |  |  | 63 | 322 | 957 |  | 6708 | 11791 | 1997 |
| 8 |  |  | 146 | 4148 |  |  | 13 | 20 | 471 |  | 1280 | 7534 | 7714 |
| 9 |  |  |  | 102 |  |  | 140 | 1 | 0 |  | 434 | 1912 | 4240 |
| 10 |  |  |  |  |  |  | 35 | 124 | 0 |  | 130 | 568 | 553 |
| 11 |  |  |  |  |  |  | 182 | 63 | 165 |  | 39 | 132 | 71 |
| 12 |  |  |  |  |  |  |  | 2573 | 0 |  | 0 | 0 | 3 |
| 13 |  |  |  |  |  |  |  |  | 2024 |  | 175 | 0 | 0 |
| 14 |  |  |  |  |  |  |  |  |  |  | 0 | 392 | 6 |
| $15+$ |  |  |  |  |  |  |  |  |  |  | 804 | 437 | 361 |
| Total | 7408 | 6166 | 5462 | 4895 | - | - | 4742 | 18474 | 25756 | - | 22496 | 28840 | 19903 |

In 1992, 1993 and 1997 there was no estimate due to poor weather conditions.
No surveys have been conducted after 2000 .
Table 3.2.3.2 Norwegian Spring-spawning herring. Estimates obtained on the acoustic surveys in the wintering areas in November-December. Numbers in millions.

| Year <br> Age | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2 *}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 |  |  |  |  |  |  | 380 |  | 9 | 65 | 74 |
| 2 | 36 | 1518 | 16 | 183 | 1465 | 73 | 1207 | 159 | 322 | 362 | 7 |
| 3 | 1247 | 2389 | 3708 | 5133 | 3008 | 661 | 441 | 2425 | 1522 | 3916 | 276 |
| 4 | 1317 | 3287 | 4124 | 5274 | 13180 | 1480 | 1833 | 296 | 5260 | 1528 | 1659 |
| 5 | 173 | 1267 | 2593 | 1839 | 5637 | 6110 | 3869 | 837 | 165 | 2615 | 624 |
| 6 | 16 | 13 | 1096 | 1040 | 994 | 4458 | 12052 | 2066 | 497 | 82 | 1029 |
| 7 | 208 | 13 | 34 | 308 | 552 | 1843 | 8242 | 6601 | 1869 | 338 | 32 |
| 8 | 139 | 158 | 25 | 19 | 92 | 743 | 2068 | 4168 | 4785 | 864 | 188 |
| 9 | 3742 | 26 | 196 | 13 | 0 | 66 | 629 | 755 | 3635 | 3160 | 516 |
| 10 | 69 | 4435 | 29 | 111 | 7 | 0 | 111 | 212 | 668 | 2216 | 1831 |
| 11 |  |  | 3239 | 39 | 41 | 0 | 14 | 0 | 205 | 384 | 911 |
| 12 |  |  |  | 907 | 15 | 126 | 0 | 15 | 0 | 127 | 184 |
| 13 |  |  |  |  | 393 | 0 | 392 | 0 | 0 | 0 | 0 |
| $14+$ |  |  |  |  |  | 842 | 221 | 146 | 168 | 18 | 0 |
| Total | 6947 | 13178 | 15209 | 15246 | 25384 | 16411 | 31144 | 17754 | 19152 | 16132 | 7345 |
| " |  |  |  |  |  |  |  |  |  |  |  |

*Much of the youngest yearclasses herring wintered outside the fjords this winter.
Table 3.2.3.3 Norwegian spring-spawning herring. Estimates from the international acoustic surveys on the feeding areas in the Norwegian Sea in May. Numbers in millions.

| Year <br> Age | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3 | 4114 | 1169 | 367 | 2191 | 1353 | 8312 | 6343 |
| 4 | 22461 | 3599 | 1099 | 322 | 2783 | 1430 | 9619 |
| 5 | 13244 | 18867 | 4410 | 965 | 92 | 1463 | 1418 |
| 6 | 4916 | 13546 | 16378 | 3067 | 384 | 179 | 779 |
| 7 | 2045 | 2473 | 10160 | 11763 | 1302 | 204 | 375 |
| 8 | 424 | 1771 | 2059 | 6077 | 7194 | 3215 | 847 |
| 9 | 14 | 178 | 804 | 853 | 5344 | 5433 | 1941 |
| 10 | 7 | 77 | 183 | 258 | 1689 | 1220 | 2500 |
| 11 | 155 | 288 | 0 | 5 | 271 | 94 | 1423 |
| 12 | 0 | 415 | 0 | 14 | 0 | 178 | 61 |
| 13 | 3134 | 60 | 112 | 0 | 114 | 0 | 78 |
| 14 |  | 2472 | 0 | 158 | 0 | 0 | 28 |
| $15+$ |  |  | 415 | 128 | 1135 | 85 | 26 |
| Total | 50504 | 44915 | 35987 | 25801 | 21661 | 21810 | 25438 |

Table 3.2.3.4 Norwegian Spring-spawning herring. Estimates obtained on the acoustic surveys in the wintering areas in January. Numbers in millions. No surveys carried out in 2000-2002.

| Year <br> Age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 90 |  |  |  | 73 |  |  |  | 214 |
| 3 | 220 | 410 | 61 | 642 | 47 | 315 |  | 267 | 1358 |
| 4 | 70 | 820 | 1905 | 3431 | 3781 | 10442 |  | 1938 | 199 |
| 5 | 20 | 260 | 2048 | 4847 | 4013 | 13557 |  | 4162 | 1455 |
| 6 | 180 | 60 | 256 | 1503 | 2445 | 4312 |  | 9647 | 4452 |
| 7 | 150 | 510 | 27 | 102 | 1215 | 1271 |  | 6974 | 12971 |
| 8 | 5500 | 120 | 269 | 29 | 42 | 290 |  | 1518 | 7226 |
| 9 | 440 | 4690 | 182 | 161 | 24 | 22 |  | 743 | 1876 |
| 10 |  | 30 | 5691 | 131 | 267 | 25 |  | 16 | 499 |
| 11 |  |  | 128 | 3679 | 29 | 200 |  | 4 | 16 |
| 12 |  |  |  |  | 4326 | 58 |  | 0 | 16 |
| 13 |  |  |  |  |  | 1146 |  | 181 | 0 |
| 14 |  |  |  |  |  |  |  | 7 | 156 |
| $15+$ |  |  |  |  |  |  |  | 314 | 220 |
| Total | 6670 | 6900 | 10567 | 14598 | 16189 | 31638 | - | 25985 | 30444 |

In 1997 there was no estimate due to poor weather conditions.
Table 3.2.3.4.1 Norwegian spring-spawning herring. Acoustic estimates (billion individuals) of immature herring in the Barents Sea in May/June. 1990-2002.See footnotes.

| Year | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}^{1}$ | $\mathbf{1 9 9 7}^{2}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 24.3 | 32.6 | 102.7 | 6.6 | 0.5 | 0.1 | 2.6 | 9.5 | 49.5 | 105.4 | 0.3 | 0.5 |
| 2 | 5.2 | 14.0 | 25.8 | 59.2 | 7.7 | 0.25 | 0.04 | 4.7 | 4.9 | 27.9 | 7.6 | 3.9 |
| 3 |  | 5.7 | 1.5 | 18.0 | 8.0 | 1.8 | 0.4 | 0.01 | 0.00 | 0.00 | 8.8 | 0.0 |
| 4 |  |  |  | 1.7 | 1.1 | 0.6 | 0.35 | 0.01 | 0.00 | 0.00 | 0.00 | 0.0 |
| 5 |  |  |  |  |  | 0.03 | 0.05 | 0.00 | 0.00 | 0.00 | 0.00 | 0.0 |

${ }^{1}$ Average of Norwegian and Russian estimates
${ }^{2}$ Combination of Norwegian and Russian estimates as described in 1998 WG report, since then only Russian estimates

Table 3.2.3.4.2 Norwegian spring-spawning herring. Acoustic estimates (billion individuals) of immature herring in the Barents Sea in September/October 2000-2001 and October/December 2002.

| Year <br> Age | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| :--- | :---: | ---: | ---: |
| 0 | 0.00 | 0.00 | 106.5 |
| 1 | 14.7 | 0.5 | 1,3 |
| 2 | 11.5 | 10.5 | 0.00 |
| 3 | 0.00 | 1.7 | 0.00 |
| 4 | 0.00 | 0.00 | 0.00 |
| 5 | 0.00 | 0.00 | 0.00 |

Table 3.2.3.4.3. Norwegian spring spawners. Acoustic abundance ( $\mathrm{TS}=20 \operatorname{logL}-71.9$ ) of 0 -group herring in Norwegian coastal waters in 1975-2002 (numbers in millions).

| Year | Area |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | South of $62^{\circ} \mathrm{N}$ | $62^{\circ} \mathrm{N}-65^{\circ} \mathrm{N}$ | $65^{\circ} \mathrm{N}-68^{\circ} \mathrm{N}$ | North of $68^{\circ} 30^{\prime}$ |  |
| 1975 |  | 164 | 346 | 28 | 538 |
| 1976 |  | 208 | 1305 | 375 | 1888 |
| 1977 |  | 35 | 153 | 19 | 207 |
| 1978 |  | 151 | 256 | 196 | 603 |
| 1979 |  | 455 | 1130 | 144 | 1729 |
| 1980 |  | 6 | 2 | 109 | 117 |
| 1981 |  | 132 | 1 | 1 | 134 |
| 1982 |  | 32 | 286 | 1151 | 1469 |
| 1983 |  | 162 | 2276 | 4432 | 6866 |
| 1984 |  | 2 | 234 | 465 | 701 |
| 1985 |  | 221 | 177 | 104 | 502 |
| 1986 |  | 5 | 72 | 127 | 204 |
| 1987 |  | 327 | 26 | 57 | 410 |
| 1988 |  | 14 | 552 | 708 | 1274 |
| 1989 |  | 575 | 263 | 2052 | 2890 |
| 1990 |  | 75 | 146 | 788 | 1009 |
| 1991 |  | 80 | 299 | 2428 | 2807 |
| 1992 |  | 73 | 1993 | 621 | 2891 |
| 1993 | 290 | 109 | 140 | 288 | 827 |
| 1994 | 157 | 452 | 323 | 6168 | 7101 |
| 1995 | 0 | 27 | 2 | 0 | 29 |
| 1996 | 0 | 20 | 114 | 8800 | 8934 |
| 1997 | 208 | 69 | 544 | 5244 | 6065 |
| 1998 | 424 | 273 | 442 | 11640 | 12779 |
| 1999 | 121 | 658 | 271 | 6329 | 7379 |
| 2000 | 570 | 127 | 996 | 7237 | 8930 |
| 2001 | 89 | 324 | 134 | 1421 | 1968 |
| 2002 | 67 | 1227 | 284 | 3573 | 5151 |

Table 3.2.3.4.4. Norwegian spring-spawning herring. Abundance indices for 0-group herring in the Barents Sea, 1974-2002.

| Year | Log index | Year | Log index |
| ---: | ---: | ---: | ---: |
| 1974 | 0.01 | 1989 | 0.58 |
| 1975 | 0.00 | 1990 | 0.31 |
| 1976 | 0.00 | 1991 | 1.19 |
| 1977 | 0.01 | 1992 | 1.05 |
| 1978 | 0.02 | 1993 | 0.75 |
| 1979 | 0.09 | 1994 | 0.28 |
| 1980 | 0.00 | 1995 | 0.16 |
| 1981 | 0.00 | 1996 | 0.65 |
| 1982 | 0.00 | 1997 | 0.39 |
| 1983 | 1.77 | 1998 | 0.59 |
| 1984 | 0.34 | 1999 | 0.41 |
| 1985 | 0.23 | 2000 | 0.30 |
| 1986 | 0.00 | 2001 | 0.13 |
| 1987 | 0.00 | 2002 | 0.53 |
| 1988 | 0.30 |  |  |
|  |  |  |  |

Table 3.2.3.5. The indices for herring larvae for the period 1981-2003 ( $\mathrm{N} * 10^{-12}$ )

| Year | Index 1 | Index 2 | Year | Index 1 | Index 2 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1981 | 0.3 |  | 1992 | 6.3 | 27.8 |
| 1982 | 0.7 |  | 1993 | 24.7 | 78.0 |
| 1983 | 2.5 |  | 1994 | 19.5 | 48.6 |
| 1984 | 1.4 |  | 1995 | 18.2 | 36.3 |
| 1985 | 2.3 |  | 1996 | 27.7 | 81.7 |
| 1986 | 1.0 |  | 1997 | 66.6 | 147.5 |
| 1987 | 1.3 | 4.0 | 1998 | 42.4 | 138.6 |
| 1988 | 9.2 | 25.5 | 1999 | 19.9 | 73.0 |
| 1989 | 13.4 | 28.7 | 2000 | 19.8 | 127.5 |
| 1990 | 18.3 | 29.2 | 2001 | 40.7 |  |
| 1991 | 8.6 | 23.5 | 2002 | 27.1 |  |
|  |  |  | 2003 | 3.7 |  |
|  |  |  |  |  |  |

Table 3.2.4.1 Tagging data used in the SeaStar runs. Please note that the tagging data for 2002 was considered an outlier and thus not included in the analysis.

Tagging data for the 1983 year class

|  | $\begin{aligned} & \underline{Z} \\ & \frac{1}{3} \\ & \overline{0} \\ & \text { D } \end{aligned}$ |  | Recovered |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  | $\begin{array}{\|l} \hline \stackrel{\rightharpoonup}{0} \\ \varrho \\ N \\ \stackrel{\rightharpoonup}{D} \\ \bar{D} \\ \\ \hline \end{array}$ |  |  |  |  |  |  |
| 1987 | 33067 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1988 | 38152 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1989 | 20620 | 10695 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 12 |
| 1990 | 24585 | 5489 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 4 |
| 1991 | 12558 | 5545 |  |  |  |  |  |  |  |  |  |  |  |  | 5 | 7 |  |
| 1992 | 15262 | 1737 |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 0 | 4 |
| 1993 | 15839 | 9372 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 6 |
| 1994 | 5364 | 9474 |  |  |  |  |  |  |  |  | 1 |  |  | 8 | 7 |  | 2 |
| 1995 | 859 | 11554 |  |  |  |  |  |  |  | 7 |  |  | 6 |  | 5 |  | 6 |
| 1996 | 2879 | 4038 |  |  |  |  |  |  | 3 | 4 |  |  | 2 |  | 6 | 2 | 3 |
| 1997 | 2266 | 3867 |  |  |  |  |  | 0 | 0 | 0 |  |  |  | 3 | 2 | 3 | 0 |
| 1998 | 648 | 509 |  |  |  |  | 1 | 0 | 0 | 0 |  |  |  |  | 1 | 3 |  |
| 1999 |  | 379 |  |  |  | 1 | 0 | 0 | 0 | 1 |  |  | 0 |  | 1 | 0 | 0 |
| 2000 |  | 413 |  |  | 0 | 0 | 0 | 0 | 0 | 0 |  |  | 0 | 3 | 0 |  | 0 |
| 2001 |  | 35 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  | 0 |  | 0 |
| 2002 |  | 221 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | - | 0 |  | 0 | 1 |  |



Table 3.2.4.1 continued. Tagging data for the 1985 year class

*1985+ group
Tagging data for the 1986 year class


Table 3.2.4.1 continued

*1987+group

Tagging data for the 1988 year class


Table 3．2．4．1 continued

| $\begin{aligned} & \text { § } \\ & \text { D } \end{aligned}$ | 2 <br> 3 <br> 3 | $\begin{aligned} & \text { Z } \\ & \bar{y} \\ & \bar{\square} \\ & \hline \end{aligned}$ | Recovered |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { D } \\ & \text { D } \\ & \text { O} \\ & \text { © } \\ & \hline \end{aligned}$ | $\begin{aligned} & \stackrel{\rightharpoonup}{0} \\ & 0 \\ & \stackrel{0}{\circ} \\ & \stackrel{2}{2} \end{aligned}$ |  | $\begin{gathered} 2000 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1998 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1997 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1996 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1995 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1994 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1993 \\ \text { release } \end{gathered}$ |
| 1993 | 7584 |  |  |  |  |  |  |  |  |
| 1994 | 11873 |  |  |  |  |  |  |  |  |
| 1995 | 2348 | 9463 |  |  |  |  |  |  | 4 |
| 1996 | 5170 | 4636 |  |  |  |  |  | 5 |  |
| 1997 | 4103 | 3346 |  |  |  |  | 0 | 7 | 2 |
| 1998 | 1176 | 1183 |  |  |  | 1 | 0 | 0 | 0 |
| 1999 |  | 1179 |  |  | 1 | 1 | 0 | 0 | 1 |
| 2000 | 470 | 790 |  | 1 | 0 | 0 | 0 | 2 | 0 |
| 2001 |  | 841 |  | 0 | 0 | 2 | 0 | 1 | 1 |
| 2002 | 319 | 286 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |

Tagging data for the 1990 year class

| $\begin{aligned} & \text { § } \\ & { } } \end{aligned}$ | $\begin{array}{r} \text { Z } \\ -\underset{y}{3} \end{array}$ |  | Recovered |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { D } \\ & \text { D } \\ & \text { O} \\ & \text { © } \\ & \hline \end{aligned}$ | $\begin{aligned} & \stackrel{\rightharpoonup}{0} \\ & \stackrel{O}{\circ} \\ & \stackrel{\otimes}{2} \end{aligned}$ |  | $\begin{array}{\|c\|} \hline 2000 \\ \text { release } \end{array}$ | $\begin{gathered} 1998 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1997 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1996 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1995 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1994 \\ \text { release } \end{gathered}$ | $\left\lvert\, \begin{gathered} 1993 \\ \text { release } \end{gathered}\right.$ |
| 1994 | 10784 |  |  |  |  |  |  | 0 |  |
| 1995 | 3868 |  |  |  |  |  | 0 | 3 |  |
| 1996 | 6171 | 9009 |  |  |  | 3 | 3 | 9 |  |
| 1997 | 4057 | 9830 |  |  | 2 | 3 | 3 | 7 |  |
| 1998 | 2381 | 2828 |  | 2 | 3 | 1 | 1 | 1 |  |
| 1999 |  | 3402 |  | 3 | 1 | 2 | 2 | 1 |  |
| 2000 | 1219 | 3146 |  | 1 | 0 | 2 | 2 | 0 |  |
| 2001 |  | 1052 |  | 0 | 0 | 0 | 0 | 2 |  |
| 2002 | 1605 | 1348 | 0 | 1 | 0 | 1 | 0 | 0 |  |

Tagging data for the 1991 year class

| $\begin{aligned} & \text { § } \\ & \stackrel{1}{7} \end{aligned}$ | $\begin{aligned} & \underline{Z} \\ & ⿳ 亠 二 口 斤 彡 ~ \end{aligned}$ | 2 $\frac{3}{3}$ $\frac{3}{0}$ | Recovered |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { D } \\ & \text { D } \\ & \stackrel{\text { O}}{\infty} \\ & \text { © } \end{aligned}$ | $\begin{aligned} & \stackrel{\rightharpoonup}{0} \\ & \stackrel{0}{\circ} \\ & \stackrel{\circ}{2} \end{aligned}$ |  | $\begin{array}{\|c\|} \hline 2000 \\ \text { release } \end{array}$ | $\begin{gathered} 1998 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1997 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1996 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1995 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1994 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1993 \\ \text { release } \end{gathered}$ |
| 1995 | 21528 |  |  |  |  |  |  |  |  |
| 1996 | 25683 |  |  |  |  |  |  |  |  |
| 1997 | 7129 | 30952 |  |  |  |  | 21 |  |  |
| 1998 | 6002 | 12459 |  |  |  | 6 | 8 |  |  |
| 1999 |  | 14968 |  |  | 4 | 14 | 7 |  |  |
| 2000 | 3802 | 18461 |  | 9 | 1 | 10 | 7 |  |  |
| 2001 |  | 10032 |  | 1 | 2 | 5 | 3 |  |  |
| 2002 | 5878 | 8937 | 10 | 9 | 1 | 1 | 1 |  |  |

Table 3.2.4.1 continued Tagging data for the 1992 year class


Tagging data for the 1993 year class


Tagging data for the 1994 year class

| $\begin{aligned} & \text { § } \\ & \text { ®1 } \end{aligned}$ | $\begin{aligned} & \text { Z } \\ & \vdots \\ & 3 \\ & 0 \end{aligned}$ | 2 $\vdots$ 3 0 <br> z | Recovered |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & \text { シ } \\ & 00 \\ & \stackrel{0}{2} \\ & \hline \end{aligned}$ |  | $\begin{gathered} 2000 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1998 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1997 \\ \text { release } \end{gathered}$ | $\begin{aligned} & 1996 \\ & \text { release } \end{aligned}$ |
| 1998 1999 2000 2001 2002 | $\begin{aligned} & 3752 \\ & 2278 \\ & 1143 \end{aligned}$ | $\begin{aligned} & 2450 \\ & 1104 \\ & 1588 \end{aligned}$ |  |  |  |  |

Tagging data for the 1995 year class

| $\begin{aligned} & \text { § } \\ & \text { 刃 } \end{aligned}$ | $\begin{aligned} & \underline{Z} \\ & 3 \\ & 3 \end{aligned}$ | 2 <br> 3 <br> 3 | Recovered |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { DO } \\ & \text { D } \\ & \stackrel{\text { O}}{\gtrless} \\ & \stackrel{\text { D }}{2} \end{aligned}$ | $\begin{aligned} & \stackrel{\stackrel{0}{0}}{0} \\ & \stackrel{\circ}{\circ} \\ & \hline \end{aligned}$ |  | $\begin{gathered} 2000 \\ \text { release } \end{gathered}$ | $\begin{aligned} & 1998 \\ & \text { release } \end{aligned}$ | $\begin{aligned} & 1997 \\ & \text { release } \end{aligned}$ | $\begin{aligned} & 1996 \\ & \text { release } \end{aligned}$ |
| $\begin{aligned} & \hline 2000 \\ & 2001 \\ & 2002 \\ & \hline \end{aligned}$ | $\begin{array}{r} 505 \\ 197 \\ \hline \end{array}$ | $\begin{aligned} & 276 \\ & 250 \end{aligned}$ | 1 |  |  |  |

Table 3.3.1.1. Estimated parameters, spawning stock biomasses in 2003, and the contribution to the likelihood function from the various data sources for the exploratory runs.

|  | Exploratory run |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Default | arvae | Youngest1998 | EstimateMortality | Lognormal | CorrectAgeBias |
| Spawning stock | 5.78 | 6.346 | 4.686 | 5.199 | 5.52 | 6.279 |
| terminalF83 | 0.189 | 0.171 | 0.189 | 0.208 | 0.238 | 0.165 |
| terminalF90 | 0.129 | 0.134 | 0.13 | 0.148 | 0.18 | 0.102 |
| terminalF91 | 0.189 | 0.168 | 0.188 | 0.211 | 0.199 | 0.161 |
| terminalF92 | 0.214 | 0.173 | 0.211 | 0.235 | 0.215 | 0.187 |
| terminalF93 | 0.216 | 0.208 | 0.213 | 0.235 | 0.228 | 0.201 |
| terminalF96 | 0.177 | 0.179 | 0.184 | 0.196 | 0.202 | 0.189 |
| terminalF97 | 0.057 | 0.054 | 0.226 | 0.065 | 0.057 | 0.055 |
| terminalF98 | 0.062 | 0.057 | 0.072 | 0.07 | 0.063 | 0.058 |
| terminalF99 | 0.021 | 0.02 | 0.022 | 0.025 | 0.022 | 0.02 |
| terminalF00 | 0.033 | 0.033 | 0.046 | 0.042 | 0.032 | 0.033 |
| terminalF01 | 0 | 0 | 0 | 0 | 0 | 0 |
| terminalF02 | 0 | 0 | 0 | 0 | 0 | 0 |
| cat1 | 0.839 | 0.8 | 0.837 | 0.924 | 0.853 | 0.784 |
| cat2 | 0.769 | 0.733 | 0.767 | 0.853 | 0.74 | 0.719 |
| cat3 | 0.844 | 0.816 | 0.846 | 0.931 | 0.853 | 0.79 |
| cat5 | 0.995 | 0.944 | 0.974 | 1.102 | 0.912 | 0.948 |
| catLarvae2 | 4.165 | 4.187 | 4.538 | 4.354 | 3.905 |  |
| cat4 | 0.427 | 0.412 | 0.518 | 0.731 | 0.436 | 0.415 |
| cat6 | 0.309 | 0.293 | 0.402 | 0.457 | 0.314 | 0.296 |
| catZero | 0.003 | 0.003 | 0.003 | 0.013 | 0.003 | 0.003 |
| Log-likelihood per term, survey 1 | -1.519 | -1.487 | -1.514 | -1.521 | -1.552 | -1.489 |
| Log-likelihood per term, survey 2 | -1.324 | -1.333 | -1.284 | -1.334 | -1.361 | -1.328 |
| Log-likelihood per term, survey 3 | -1.646 | -1.653 | -1.64 | -1.658 | -1.639 | -1.607 |
| Log-likelihood per term, survey 4 | -3.494 | -3.485 | -3.511 | -3.477 | -3.528 | -3.486 |
| Log-likelihood per term, survey 5 | -1.899 | -1.882 | -1.833 | -1.888 | -1.902 | -1.899 |
| Log-likelihood per term, survey 6 | -3.101 | -3.09 | -3.078 | -3.118 | -3.142 | -3.094 |
| Number of data points, survey 1 | 21 | 21 | 21 | 21 | 21 | 21 |
| Number of data points, survey 2 | 41 | 41 | 43 | 41 | 41 | 41 |
| Number of data points, survey 3 | 19 | 19 | 19 | 19 | 19 | 19 |
| Number of data points, survey 4 | 19 | 19 | 19 | 19 | 19 | 19 |
| Number of data points, survey 5 | 36 | 36 | 41 | 36 | 36 | 36 |
| Number of data points, survey 6 | 7 | 7 | 7 | 7 | 7 | 7 |
| vpaM | 0.15 | 0.15 | 0.15 | 0.128 | 0.15 | 0.15 |
| vpaMYoung | 0.9 | 0.9 | 0.9 | 0.507 | 0.9 | 0.9 |

Table 3.3.1.2. Estimated parameters, spawning stock biomasses in 2003, and the contribution to the likelihc function from the various data sources for the exploratory runs including only one survey

|  | Exploratory run |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | OnlySurvey1 | OnlySurvey2 | OnlySurvey3 | OnlySurvey5 |
| Spawning stock | 7.771 | 6.314 | 6.921 | 4.757 |
| terminalF83 | 0.153 | 0.177 | 0.352 | 0.221 |
| terminalF90 | 0.066 | 0.091 | 0.096 | 0.167 |
| terminalF91 | 0.167 | 0.171 | 0.245 | 0.258 |
| terminalF92 | 0.237 | 0.229 | 0.321 | 0.284 |
| terminalF93 | 0.215 | 0.225 | 0.157 | 0.194 |
| terminalF96 | -0.042 | 0.093 | 0.044 | 0.378 |
| terminalF97 | 0.05 | 0.055 | 0.052 | 0.064 |
| terminalF98 | 0.064 | 0.064 | 0.071 | 0.067 |
| terminalF99 | 0.022 | 0.022 | 0.025 | 0.023 |
| terminalF00 | -0.011 | 0.023 | 0.009 | 0.038 |
| terminalF01 | 0 | 0 | 0 | 0 |
| terminalF02 | 0 | 0 | 0 | 0 |
| cat1 | 0.753 |  |  |  |
| catLarvae2 | 3.528 | 3.866 | 3.882 | 4.583 |
| cat4 | 0.312 | 0.382 | 0.309 | 0.486 |
| cat6 | 0.263 | 0.299 | 0.263 | 0.332 |
| catZero | 0.003 | 0.003 | 0.004 | 0.004 |
| cat2 |  | 0.694 |  |  |
| cat3 |  |  | 0.908 |  |
| cat5 |  |  |  | 1.112 |
| Log-likelihood per term, survey 1 | -1.558 |  |  |  |
| Log-likelihood per term, survey 2 |  | -1.299 |  |  |
| Log-likelihood per term, survey 3 |  |  | -1.402 |  |
| Log-likelihood per term, survey 4 | -3.603 | -3.533 | -3.643 | -3.468 |
| Log-likelihood per term, survey 5 |  |  |  | -1.764 |
| Log-likelihood per term, survey 6 | -3.199 | -3.132 | -3.25 | -3.096 |
| Number of data points, survey 1 | 21 | 0 | 0 | 0 |
| Number of data points, survey 2 | 0 | 41 | 0 | 0 |
| Number of data points, survey 3 | 0 | 0 | 19 | 0 |
| Number of data points, survey 4 | 19 | 19 | 19 | 19 |
| Number of data points, survey 5 | 0 | 0 | 0 | 36 |
| Number of data points, survey 6 | 7 | 7 | 7 | 7 |
| vpaM | 0.15 | 0.15 | 0.15 | 0.15 |
| vpaMYoung | 0.9 | 0.9 | 0.9 | 0.9 |

Table 3.3.2. Summary of the bootstrap runs

|  | Short name | ML value | Mean | Median | Standard deviation |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Spawning stock, assesment year | SSB | 5.752 | 5.456 | 5.31 | 0.929 |
| Total stock, assessment year | TSB | 7.849 | 7.828 | 7.827 | 1.363 |
| distribution Parameter | d | 0.457 | 0.481 | 0.481 | 0.002 |
| cat1 | c1 | 0.839 | 0.884 | 0.886 | 0.019 |
| cat2 | c2 | 0.769 | 0.79 | 0.791 | 0.023 |
| cat3 | c3 | 0.844 | 0.883 | 0.883 | 0.014 |
| cat5 | c5 | 0.995 | 0.964 | 0.963 | 0.026 |
| catLarvae2 | cL2 | 4.166 | 4.4 | 4.32 | 0.825 |
| distributionParameterLarvae | dL | 0.692 | 0.923 | 0.932 | 0.098 |
| terminalF83 | F83 | 0.19 | 0.254 | 0.255 | 0.009 |
| terminalF90 | F90 | 0.129 | 0.202 | 0.201 | 0.011 |
| terminalF91 | F91 | 0.189 | 0.219 | 0.219 | 0.011 |
| terminalF92 | F92 | 0.214 | 0.236 | 0.236 | 0.012 |
| terminalF93 | F93 | 0.217 | 0.249 | 0.249 | 0.01 |
| terminalF96 | F96 | 0.177 | 0.224 | 0.223 | 0.014 |
| taggingSurvival | tS | 0.394 | 0.376 | 0.376 | 0.006 |
| distributionParameter | d | 1.942 | 1.856 | 1.809 | 0.072 |
| cat4 | c4 | 0.423 | 0.457 | 0.454 | 0.062 |
| cat6 | c6 | 0.309 | 0.333 | 0.324 | 0.076 |
| catZero | cZ | 0.003 | 0.003 | 0.003 | 0 |
| distParZero | dZ | 0.21 | 0.143 | 0.141 | 0.046 |
| terminalF97 | F97 | 0.056 | 0.09 | 0.063 | 0.077 |
| terminalF98 | F98 | 0.063 | 0.064 | 0.059 | 0.024 |
| terminalF99 | F99 | 0.022 | 0.023 | 0.02 | 0.01 |
| terminalF00 | F00 | 0.03 | 0.047 | 0.048 | 0.024 |
| terminalF01 | F01 | 0 | 0 | 0 | 0 |
| terminalF02 | F02 | 0 | $*$ | 0 | $*$ |
| nostimabl |  |  |  |  |  |

* not estimable

Table 3.3.3 Summary table SeaStar retrospective run, years 1950-2002. Recruits are billion, biomass are million tonnes, landings are tonnes.

| Year | Recruits Age 0 | Total biomass | Spawning stock biomass | Landings | Fbar 5-14 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 693.066 | 17.383 | 12.955 | 933000 | 0.06 |
| 1951 | 143.606 | 17.43 | 11.964 | 1278400 | 0.073 |
| 1952 | 96.42 | 18.496 | 10.706 | 1254800 | 0.078 |
| 1953 | 86.254 | 16.558 | 9.616 | 1090600 | 0.073 |
| 1954 | 44.367 | 18.087 | 9.256 | 1644500 | 0.129 |
| 1955 | 25.473 | 15.605 | 9.848 | 1359800 | 0.088 |
| 1956 | 30.646 | 13.902 | 11.43 | 1659400 | 0.124 |
| 1957 | 25.397 | 11.313 | 10.129 | 1319500 | 0.126 |
| 1958 | 23.095 | 9.799 | 9.157 | 986600 | 0.096 |
| 1959 | 412.475 | 8.406 | 7.694 | 1111100 | 0.137 |
| 1960 | 197.51 | 8.084 | 6.457 | 1101800 | 0.166 |
| 1961 | 76.099 | 8.264 | 4.979 | 830100 | 0.126 |
| 1962 | 19.006 | 7.238 | 4.019 | 848600 | 0.172 |
| 1963 | 168.933 | 7.457 | 3.291 | 984500 | 0.299 |
| 1964 | 93.902 | 7.081 | 3.232 | 1281800 | 0.241 |
| 1965 | 8.49 | 6.443 | 3.672 | 1547700 | 0.277 |
| 1966 | 51.405 | 5.065 | 3.561 | 1955000 | 0.69 |
| 1967 | 3.943 | 3.587 | 2.084 | 1677200 | 1.496 |
| 1968 | 5.184 | 1.45 | 0.823 | 712200 | 3.414 |
| 1969 | 9.276 | 0.634 | 0.596 | 67800 | 0.547 |
| 1970 | 0.661 | 0.508 | 0.469 | 62300 | 1.211 |
| 1971 | 0.24 | 0.471 | 0.381 | 21100 | 1.554 |
| 1972 | 2.376 | 0.386 | 0.322 | 13161 | 1.689 |
| 1973 | 13.629 | 0.452 | 0.422 | 7017 | 1.645 |
| 1974 | 8.631 | 0.461 | 0.371 | 7619 | 0.13 |
| 1975 | 3.027 | 0.569 | 0.33 | 13713 | 0.217 |
| 1976 | 8.08 | 0.597 | 0.372 | 10436 | 0.124 |
| 1977 | 5.092 | 0.636 | 0.506 | 22706 | 0.077 |
| 1978 | 6.2 | 0.745 | 0.561 | 19824 | 0.039 |
| 1979 | 12.035 | 0.783 | 0.567 | 12864 | 0.022 |
| 1980 | 1.473 | 0.865 | 0.618 | 18577 | 0.032 |
| 1981 | 1.1 | 0.887 | 0.617 | 13736 | 0.022 |
| 1982 | 2.329 | 0.806 | 0.598 | 16655 | 0.021 |
| 1983 | 324.67 | 1.149 | 0.653 | 23054 | 0.03 |
| 1984 | 11.528 | 2.061 | 0.661 | 53532 | 0.093 |
| 1985 | 35.62 | 5.237 | 0.568 | 169872 | 0.393 |
| 1986 | 6.041 | 1.856 | 0.472 | 225256 | 1.139 |
| 1987 | 9.011 | 3.166 | 0.928 | 127306 | 0.447 |
| 1988 | 27.852 | 3.535 | 2.835 | 135301 | 0.045 |
| 1989 | 71.251 | 4.139 | 3.441 | 103830 | 0.028 |
| 1990 | 126.958 | 4.652 | 3.601 | 86411 | 0.022 |
| 1991 | 335.96 | 5.36 | 3.743 | 84683 | 0.025 |
| 1992 | 377.724 | 7.153 | 3.606 | 104448 | 0.029 |
| 1993 | 99.413 | 7.274 | 3.51 | 232457 | 0.066 |
| 1994 | 32.509 | 8.167 | 4.011 | 479228 | 0.134 |
| 1995 | 8.738 | 7.803 | 4.261 | 905501 | 0.222 |
| 1996 | 72.523 | 8.011 | 4.106 | 1220283 | 0.183 |
| 1997 | 102.981 | 8.243 | 7.094 | 1426507 | 0.176 |
| 1998 | 201.232 | 8.152 | 7.222 | 1223131 | 0.156 |
| 1999 | 150.176 | 8.199 | 6.734 | 1235433 | 0.194 |
| 2000 | 17.561 | 8.526 | 5.443 | 1207201 | 0.236 |
| 2001 | 5.006 | 7.473 | 4.903 | 770054 | 0.182 |
| 2002 | 158.994 | 7.595 | 5.226 | 806086 | 0.153 |

Table 3.3.4.1 Input data to the short term projection based on SeaStar.

## Landings in 2003

$$
0.71 \text { million tonnes }
$$ 7.8933 million tonnes 5.78017 million tonnes

$5-14$, Fbar is weighted with population numbers January 1
0.105151
assumed equal
predicted condition factor
mean of last 3 years

|  | Age Numbers(billion) $\chi^{\text {W }}$ Weightstock(kg) 2 Weightstock(kg) 2 Weightcatch(kg) |  |  |  | Fractionmatureatage Exploitationpattern |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0.001 | 0.001 | 0 | 0 | 0 |
| 1 | 64.642 | 0.01 | 0.01 | 0 | 0 | 0.002 |
| 2 | 0.827 | 0.055 | 0.057 | 0.086 | 0 | 0.033 |
| 3 | 1.141 | 0.098 | 0.101 | 0.156 | 0 | 0.063 |
| 4 | 8.502 | 0.159 | 0.164 | 0.213 | 0.3 | 0.09 |
| 5 | 9.336 | 0.211 | 0.218 | 0.256 | 0.9 | 0.114 |
| 6 | 3.989 | 0.272 | 0.28 | 0.275 | 1 | 0.136 |
| 7 | 1.539 | 0.305 | 0.314 | 0.3 | 1 | 0.154 |
| 8 | 0.131 | 0.292 | 0.301 | 0.312 | 1 | 0.169 |
| 9 | 0.381 | 0.331 | 0.341 | 0.327 | 1 | 0.18 |
| 10 | 1.005 | 0.337 | 0.347 | 0.345 | 1 | 0.187 |
| 11 | 2.55 | 0.347 | 0.358 | 0.36 | 1 | 0.189 |
| 12 | 1.496 | 0.356 | 0.367 | 0.374 | 1 | 0.187 |
| 13 | 0.352 | 0.381 | 0.393 | 0.394 | 1 | 0.179 |
| 14 | 0.077 | 0.414 | 0.427 | 0.4 | 1 | 0.166 |
| 15 | 0.041 | 0.425 | 0.438 | 0.424 | 1 | 0.147 |
| 16 | 0.248 | 0.441 | 0.455 | 0.421 | 1 | 0.122 |

Table 3.3.4.2 Output from the short term projection based on SeaStar.

| Fbar | SSB 2004 | Landings 2004 | Biomass 2005 | SSB 2005 |
| :---: | :---: | :---: | :---: | :---: |
| 0.071 | 6.43 | 0.483 | 7.25 | 6.004 |
| 0.086 | 6.42 | 0.575 | 7.159 | 5.909 |
| 0.1 | 6.411 | 0.667 | 7.069 | 5.816 |
| 0.114 | 6.401 | 0.757 | 6.98 | 5.724 |
| 0.128 | 6.392 | 0.846 | 6.892 | 5.634 |
| 0.143 | 6.382 | 0.933 | 6.806 | 5.545 |
| 0.157 | 6.373 | 1.02 | 6.721 | 5.458 |
| 0.171 | 6.364 | 1.105 | 6.637 | 5.372 |
| 0.186 | 6.354 | 1.189 | 6.555 | 5.288 |
| 0.2 | 6.345 | 1.272 | 6.473 | 5.205 |
| 0.214 | 6.336 | 1.354 | 6.393 | 5.123 |

Table 3.4.1
Results of stock assessment for NSS herring by means of ISVPA (ISVPA, all indexes-MDN, feeging area - with SSE)

| year | B(1+) | SSB | R(1) | F(5-14, w-d by N(a)) |
| :---: | :---: | :---: | :---: | :---: |
| 1986 | 1666065 | 414988 | 6660886 | 1.191 |
| 1987 | 2816113 | 844894 | 3496148 | 0.693 |
| 1988 | 3206061 | 2640594 | 3610493 | 0.180 |
| 1989 | 3675905 | 3177113 | 10340660 | 0.065 |
| 1990 | 4061779 | 3205289 | 27655650 | 0.067 |
| 1991 | 4505327 | 3325907 | 46207070 | 0.039 |
| 1992 | 5451299 | 3202456 | 124843600 | 0.036 |
| 1993 | 6742704 | 3126439 | 139820000 | 0.055 |
| 1994 | 7826000 | 3590663 | 42391460 | 0.074 |
| 1995 | 8678664 | 4482819 | 15162190 | 0.130 |
| 1996 | 8657097 | 5972824 | 4360407 | 0.122 |
| 1997 | 8952760 | 7101314 | 43741090 | 0.221 |
| 1998 | 7827215 | 6395750 | 35626600 | 0.190 |
| 1999 | 9820659 | 5975486 | 147787200 | 0.169 |
| 2000 | 9879879 | 5062963 | 102819500 | 0.232 |
| 2001 | 10958050 | 4590611 | 153211100 | 0.125 |
| 2002 | 12664870 | 5219457 | 0 | 0.161 |

to be substituted by 538000 (survey estimate)








$\infty$




 $1454420 \quad 412231$
 $\begin{array}{rr}6546225 & 3275768 \\ 17727840 & 5532253 \\ 19834860 & 14937070\end{array}$ $\begin{array}{rr}19834860 & 14937070 \\ 5998704 & 16409970 \\ 2019622 & 4911761\end{array}$ 2013156
643452

Table 3.4.2
NSS herring. ISVPA. Abundance at age by years

[^1]$\stackrel{\curvearrowleft}{\curvearrowleft}$
 O 0.00000 0.14811
0.00000 응 $\stackrel{\infty}{\infty} \stackrel{n}{\stackrel{0}{ }}$ 0.00000

$\stackrel{\downarrow}{\leftarrow}$

응

 0.05333
0.00000 0.11279
0.25181 す o
N
M
O
 N $\stackrel{\infty}{\stackrel{\rightharpoonup}{ \pm}} \stackrel{\underset{\sim}{\infty}}{\stackrel{\rightharpoonup}{N}}$




$\infty$


へ $\qquad$ 0.36053 0.03564 0.02695
0.03381
0.05616 $\stackrel{N}{N}$

 | N |
| :---: |
|  |
|  |
|  |

 $\underset{\sim}{\sim}$
$\bullet$


م $\qquad$ 0.26918
0.22074
0.17506 0.04426
0.03649
0.01875 0.01875
0.01421 0.01780

0.02940 0.06238 | $\pm$ | $\bar{J}$ |
| :---: | :---: |
| 0 |  | 0.11399 0.08432

0.09491


ल L
O
O
0
0 0.03842
0.03104
0.00827
0.00684
0.00354 0.00269 0.00553
0.01157 N ~~~ $\bar{\circ}$
O

0 0.01551 | $\square$ |
| :---: |
| O |
|  |

r


Table 3.4.4. NSS herring. ISVPA. Input data for short term projection

Fbar age range: 5-14

| 2003 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Stock size | Natural mortality | Maturity ogive | Prop. of F bef. spaw. | Prop. of M bef. spaw. | Weight in stock | Exploit. pattern | Weight in catch |
| 1 | 28571 | 0.9 | 0.00 | 0.2 | 0.2 | 0.018 | 0.000 | 0.018 |
| 2 | 219 | 0.9 | 0.00 | 0.2 | 0.2 | 0.054 | 0.010 | 0.086 |
| 3 | 3330 | 0.15 | 0.00 | 0.2 | 0.2 | 0.098 | 0.078 | 0.156 |
| 4 | 14444 | 0.15 | 0.30 | 0.2 | 0.2 | 0.159 | 0.208 | 0.213 |
| 5 | 17415 | 0.15 | 0.90 | 0.2 | 0.2 | 0.211 | 0.424 | 0.256 |
| 6 | 3331 | 0.15 | 1.00 | 0.2 | 0.2 | 0.272 | 0.623 | 0.275 |
| 7 | 2832 | 0.15 | 1.00 | 0.2 | 0.2 | 0.305 | 0.824 | 0.300 |
| 8 | 194 | 0.15 | 1.00 | 0.2 | 0.2 | 0.314 | 0.966 | 0.312 |
| 9 | 512 | 0.15 | 1.00 | 0.2 | 0.2 | 0.331 | 1.223 | 0.345 |
| 10 | 1120 | 0.15 | 1.00 | 0.2 | 0.2 | 0.337 | 1.757 | 0.360 |
| 11 | 1865 | 0.15 | 1.00 | 0.2 | 0.2 | 0.346 | 1.697 | 0.374 |
| 12 | 993 | 0.15 | 1.00 | 0.2 | 0.2 | 0.356 | 1.720 | 0.394 |
| 13 | 152 | 0.15 | 1.00 | 0.2 | 0.2 | 0.381 | 1.727 | 0.400 |
| 14 | 36 | 0.15 | 1.00 | 0.2 | 0.2 | 0.414 | 2.334 | 0.424 |
| 15 | 14 | 0.15 | 1.00 | 0.2 | 0.2 | 0.425 | 2.334 | 0.421 |
| 16 | 1 | 0.15 | 1.00 | 0.2 | 0.2 | 0.440 | 2.334 | 0.440 |


| 2004 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Stock <br> size | Natural mortality | Maturity ogive | Prop. of F bef. spaw. | Prop. of M bef. spaw. | Weight in stock | Exploit. pattern | Weight in catch |
| 1 | 28571 | 0.9 | 0.00 | 0.2 | 0.2 | 0.018 | 0.000 | 0.018 |
| 2 |  | 0.9 | 0.00 | 0.2 | 0.2 | 0.054 | 0.010 | 0.086 |
| 3 |  | 0.15 | 0.00 | 0.2 | 0.2 | 0.098 | 0.078 | 0.156 |
| 4 |  | 0.15 | 0.30 | 0.2 | 0.2 | 0.159 | 0.208 | 0.213 |
| 5 |  | 0.15 | 0.90 | 0.2 | 0.2 | 0.211 | 0.424 | 0.256 |
| 6 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.272 | 0.623 | 0.275 |
| 7 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.305 | 0.824 | 0.300 |
| 8 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.314 | 0.966 | 0.312 |
| 9 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.331 | 1.223 | 0.345 |
| 10 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.337 | 1.757 | 0.360 |
| 11 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.346 | 1.697 | 0.374 |
| 12 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.356 | 1.720 | 0.394 |
| 13 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.381 | 1.727 | 0.400 |
| 14 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.414 | 2.334 | 0.424 |
| 15 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.425 | 2.334 | 0.421 |
| 16 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.440 | 2.334 | 0.440 |


| 2005 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Stock size | Natural mortality | Maturity ogive | Prop. of F bef. spaw. | Prop. of M bef. spaw. | Weight in stock | Exploit. pattern | Weight in catch |
| 1 | 28571 | 0.9 | 0.00 | 0.2 | 0.2 | 0.018 | 0.000 | 0.018 |
| 2 |  | 0.9 | 0.00 | 0.2 | 0.2 | 0.054 | 0.010 | 0.086 |
| 3 |  | 0.15 | 0.00 | 0.2 | 0.2 | 0.098 | 0.078 | 0.156 |
| 4 |  | 0.15 | 0.30 | 0.2 | 0.2 | 0.159 | 0.208 | 0.213 |
| 5 |  | 0.15 | 0.90 | 0.2 | 0.2 | 0.211 | 0.424 | 0.256 |
| 6 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.272 | 0.623 | 0.275 |
| 7 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.305 | 0.824 | 0.300 |
| 8 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.314 | 0.966 | 0.312 |
| 9 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.331 | 1.223 | 0.345 |
| 10 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.337 | 1.757 | 0.360 |
| 11 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.346 | 1.697 | 0.374 |
| 12 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.356 | 1.720 | 0.394 |
| 13 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.381 | 1.727 | 0.400 |
| 14 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.414 | 2.334 | 0.424 |
| 15 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.425 | 2.334 | 0.421 |
| 16 |  | 0.15 | 1.00 | 0.2 | 0.2 | 0.440 | 2.334 | 0.440 |

Input units are millions and kg - output in kilotonnes
Table 3.4.5. NSS herring. ISVPA. Short term projection
Fbar: 5-14 weighted byabundance
Basis for 2003: F2003 = F2002; Recruitment: GM 1986-2001 = 28571 millions

| 2003 |  |  |  |  | 2004 |  |  |  |  | 2005 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 10279 | 7104 | 0.736 | 0.118 | 710 | 10784 | 8795 | 0.0 | 0.000 | 0 | 11187 | 9241 |
|  |  |  |  |  | . | 8773 | 0.1 | 0.016 | 112 | 11079 | 9136 |
|  |  |  |  |  | . | 8752 | 0.2 | 0.032 | 222 | 10972 | 9032 |
|  |  |  |  |  | . | 8730 | 0.3 | 0.048 | 331 | 10866 | 8930 |
|  |  |  |  |  | . | 8709 | 0.4 | 0.065 | 438 | 10762 | 8830 |
|  |  |  |  |  | . | 8688 | 0.5 | 0.081 | 543 | 10660 | 8731 |
|  |  |  |  |  | . | 8666 | 0.6 | 0.097 | 647 | 10559 | 8633 |
|  |  |  |  |  | . | 8645 | 0.7 | 0.113 | 749 | 10460 | 8537 |
|  |  |  |  |  | . | 8624 | 0.8 | 0.129 | 850 | 10362 | 8442 |
|  |  |  |  |  | . | 8603 | 0.9 | 0.145 | 949 | 10266 | 8349 |
|  |  |  |  |  | . | 8582 | 1.0 | 0.161 | 1047 | 10171 | 8257 |
|  |  |  |  |  | . | 8561 | 1.1 | 0.177 | 1143 | 10077 | 8166 |
|  |  |  |  |  | . | 8540 | 1.2 | 0.194 | 1238 | 9985 | 8077 |
|  |  |  |  |  | . | 8520 | 1.3 | 0.210 | 1332 | 9894 | 7989 |
|  |  |  |  |  | . | 8499 | 1.4 | 0.226 | 1424 | 9804 | 7902 |
|  |  |  |  |  | . | 8478 | 1.5 | 0.242 | 1515 | 9716 | 7817 |
|  |  |  |  |  | . | 8458 | 1.6 | 0.258 | 1605 | 9628 | 7732 |
|  |  |  |  |  | . | 8437 | 1.7 | 0.274 | 1694 | 9542 | 7649 |
|  |  |  |  |  |  | 8417 | 1.8 | 0.290 | 1781 | 9458 | 7567 |
|  |  |  |  |  |  | 8396 | 1.9 | 0.306 | 1867 | 9374 | 7487 |
|  |  |  |  |  |  | 8376 | 2.0 | 0.323 | 1952 | 9292 | 7407 |

Input units are millions and kg - output in kilotonnes
for $F(2003)=0.15 \operatorname{Catch}(2003)=886000$ tonnes

Table 3.5.2 Comparison of the 1992 and 1998 year class.
The year classes are compared by acoustic surveys, trawl indexes and estimates.

| Yearclass | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 8}$ |
| :--- | :---: | :---: |
| 0-group index (trawl survey) | 1.05 | 0.59 |
| Barents Sea, acoustic est., age 1, <br> (billion indv) | 102.7 | 49.5 |
| Barents Sea, acoustic est. age 2, <br> (billion indv) | 59.2 | 27.9 |
| Norwegain Sea, acoustic <br> estimate age 4 (billion indv) | 22.4 | 9.6 |
| Age 4 estimates WG $_{2002}$ VPA | 22.6 |  |
| Age 4 estimates ISVPA | 2003 | 20.0 |
| Age 4 estimates SeaStar | 2003 | 21.8 |



Figure 3.1.2.1 Total catches of Norwegian spring-spawning herring in 2002 by ICES rectangle. Grading of the symbols: black dots less than 300 t , open squares 300-3 000 t , and black squares $>3000 \mathrm{t}$.


Figure 3.1.2 2 Total catches of Norwegian spring-spawning herring in 2002 by quarter and ICES rectangle. Grading of the symbols: black dots less than 300 t , open squares 300-3 000 t , and black squares $>$ 3000 t .


Figure 3.2.3.5. Distribution of herring larvae April 2003


Condition (yr2) $=0.022 * N A O$ yr $1+0.822, R^{2}=0.66, P=0.007$.

Figure 3.2.4.3 Observed and modelled data of herring condition (Section 2.4.5). Prognosis of herring condition in December 2003 was based on the mean NAO for March-April in 2002:


Figure 3.3.1 Histograms of deviations from true value when large year classes are estimated separately (blue line), and when large year classes are estimated together with the small year classes (red line). See the text for details.


Figure 3.3.2.1 VPA scaled by catchability (solid line) and observations (dots) for the acoustic survey on the spawning grounds, for year classes in the tuning.


Figure 3.3.2.2 VPA scaled by catchability (solid line) and observations (dots) for the acoustic survey on the wintering areas in Vestfjorden in December, for year classes in the tuning.


Figure 3.3.2.3 VPA scaled by catchability (solid line) and observations (dots) for the acoustic survey on the wintering areas in Vestfjorden in January, for year classes in the tuning.


Figure 3.3.2.4 VPA scaled by catchability (solid line) and observations (dots) for the acoustic survey on the feeding grounds in the Norwegian Sea, for year classes in the tuning.


Figure 3.3.2.5. SeaStar retrospective plot


Figure 3.3.2.6 Histogram of spawning stock replicates from the bootstrap

|  | SSB | TSB | d | c1 | c2 | c3 | c5 | CL2 | dL | F83 | F90 | F91 | F92 | F93 | F96 | tS | d | C4 | c6 | CZ | dZ | F97 | F98 | F99 | F00 | F01 | F02 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SSB | $\cdots$ | - | \% | \% | 4 | $\square$ | \% | 4 | + | M | \% | Ime | \% | \% | \% | \% |  | \% | S | + | \% | $\cdots$ | \% | + | 4 | \# | $\pm$ |
| TSB | 4 |  | \% | \% | $\square$ | $\cdots$ | It | $\square$ | \% | F+ | \% | T+ | \% | \% | \% | \% |  | 4ta | T | 17 | 4 | $\cdots$ | $\pm$ | T+ | \% | + | $\pm$ |
| d | \% |  | $\cdots$ | \% | Mily | , ${ }^{4}$ | T | \% | H/ | V" | \% | 47 | F | , | V1/ | 4 |  | \% | - | , | 4 | $\pm$ | + | W | Stars | $\square$ | $\square$ |
| c1 | 느라․ | \% |  |  | + | \% | U | + | + | Fim | F | $\square^{*}$ | F | - ${ }^{*}$ | $\square^{*}$ | \% |  | \% | \% | \% | $\cdots$ | L | \% | $\cdots$ | + | $\square$ | $\square$ |
| c2 | \$ | M | $\square$ | 4 |  |  |  | 5 | \% | - |  | + |  |  | + | + |  | , | \# | 4 | $\pm$ | , | + | Tita |  | $\square$ | \% |
| c3 | H. | + | \% |  |  |  |  | ㄴ. | 3 |  |  | + |  |  |  | 4ter |  | 4 | $\pm$ | \# | - | 3 | + | TH |  | Y | $\cdots$ |
| c5 | + | + | T" | 4 | 4FP. |  |  | $\square$ | $3+$ |  |  | +ip | P |  | \% | \% |  | tm | \% | 4 | \% | $\pm$ | S+3 | 4 | Cle | + | $\cdots$ |
| CL2 | U2 | - | 4 | 4t | S | ] |  |  | \% | \% | \% | \% | \% | \% | \% | Tl |  | \% | 4 | IH: | $\pm$ | \% | Ha, | $\because$ | \%a | $\pm$ | $\cdots$ |
| dL | + | 4 | $\cdots$ | M | \% | 4 | $4$ | 4 |  | \% | \% | 1+3. | $\cdots$ | 4 | \% | \% |  | - | $\pm$ | + | \% | 4 | \% | $\square$ | $\cdots$ | ¢ | $\cdots$ |
| F83 | 4, | " | $\cdots$ |  | \% |  |  | 4 | " |  |  | ${ }^{+}$ |  |  | 4 | 4 |  | , | \% | I | + | $\checkmark$ |  | 4- | $\cdots$ | $\square$ | $\cdots$ |
| F90 | 4 | \# | $\square$ | W |  |  |  | \% | 4 |  |  | + | 为 |  |  | * |  | 4, | \# | \% | \% | $\bigcirc$ | 4 | \% |  | $\square$ | $\square$ |
| F91 | , | 4 | \% | - | + ${ }^{\text {F }}$ |  | +it | H | \% | 4 |  |  | \% |  | \% | 4 |  | + | T | 1 | MI | U* | W | H파는 | - | $\square$ | $\pm$ |
| F92 | \% | ¢ | F | \% | \% | + |  | M, | \% | \% |  | - |  |  | \#\# | " |  | , \% | \% | $\geqslant$ | 4 | 4 | 4 | 4 | \% | + | $\pm$ |
| F93 | V+ | + | च |  |  |  |  | 4ta | \% |  |  | + |  |  |  | $\underline{\square}$ |  | S | H+ | , | \% | 4* | 4 | $\cdots$ | F. | $\square$ | \% |
| F96 | , | 4 | $\pm$ |  |  |  |  | \% | \% |  |  | + |  |  |  | $\pm$ |  | Wt | II |  | Wif | 4 | \% | Hz | $\pm$ | + | $\pm$ |
| tS | + | 1m | \% |  | \% |  |  | \# | + | " | +4\% | \% | ㅍ. |  | ¢ |  |  |  | T | 4 | 4 | $\pm$ | + | "+ | 1 | 4 | $\cdots$ |
| d |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |  |  | .11 |  | , |  |
| C4 | TH. | + | H | $\square$ | ) | \% | 4 | $\pm$ | 4 | 17 | \% | We. | - | \# | \% | 4 |  |  | \% | 4 | , | 4 | $\square$ | 4 | Nix | \% | \% |
| c6 | + | 7 | \% | 4 | \% | " | 4 | \% | $\pm$ | \% | \% | \$ | S | 4 | W | 4 |  |  |  | \% | $\pm$ | \# | (1) | 4 | $\pm$ | $\pm$ | $\cdots$ |
| cZ | + | , | \% | M 4 | \% | , | $\pm$ | Y | , | \% | ME. | , | 3 | \% | \% | T |  | , | 4 |  | 4 | \% | It | $\because$ | $\pm$ | $\square$ | $\pm$ |
| dz | \% | 4 | $\cdots$ |  | 4 | A | \% | 4, | \% | 14 | 4 | 4 |  | \% | F | 1 | $=$ | + | $\pm$ | \% |  | 11 T | + | $\because$ | P | \% | $\cdots$ |
| F97 | + | \#+ | \#\% |  | 4 | 1 | $\pm$ | 4 | Wer | \% | + | \% |  |  | \% | \% |  |  | $\square$ | \% |  |  | + | \#\% |  | $=$ | - |
| F98 | $\square$ | 4 | $\cdots$ | + | 1 | 4 |  | +\% | \% | M | S | + | I |  |  | + |  |  | \% | 1 | \%e | 4 |  | 1 | 14 | $\pm$ | , |
| F9 | + | 4 | 4 | 3 | \% | 4 | $\pm$ | $4$ | 4 | \% | " | 4 | 4 |  | \# | \#F | $\bigcirc$ | 4 | 4 | 1 | 18 |  | 4 |  | \% | $\pm$ | $\square$ |
| F00 | \# | \% | 4 | - | + | S | F | $\pm$ | - | Fl | \% | W | - | \% | \% | , | $\cdots$ | - | 4 | \% | $\cdots$ | + | 4 | +3. |  |  | + |
| F01 | +2] | Hz | T 7 | +1 | \% | $1 \pm$ |  | W | - | 4 | \% | \% 4 | 174 | + | + + | + |  | \% | $\pm$ | \% | 4 | \#-1 | 1+ | 1] | \% | $\cdots$ | \% |
| F02 | 4 ${ }^{3}$ | Lis | \% |  | P |  |  |  | 4 | 4 | \# | 4 |  | 4 | W | $\pm \pm$ | \%"m. |  | 4 | \% |  |  |  | \% |  | 7 |  |

Figure 3.3.2.7 Scatter plots of the covariation between pairs of parameters, between parameters and spawning stock, and between parameters and total stock from SeaStar bootstrap replicates.


Figure 3.3.3 Spawning stock - recruitment plot for the SeaStar Default run. A log-based Beverton-Holt recruitment function is fitted to the least $75 \%$ successful recruitment points. The $25 \%$ most successful recruitment points (red) are treated separately during simulations


Figure 3.4.1.
NSS Herring.
ISVPA, catch-controlled version, logarithmic procedure, unbiased description of LnC(a,y)
minimization of MDN for $C(a, y)$ and SSE - for others
$q(a)$ are estimated for all indexes


Fig. 3.4.2
NSS Herring.
ISVPA, catch-controlled version, logarithmic procedure,
unbiased description of $\operatorname{LnC}(a, y)$
minimization of MDN for each kind of data
$q(a)$ are estimated for all indexes


Figure 3.4.3
NSS herring. ISVPA.
The ISVPA overall loss function for different cases with respect to survey N4


Fig. 3.4.4.
NSS herring. ISVPA
Estimates of SBB in relation to options with respect to survey N4


Figure 3.4.5
ISVPA, catch-controlled, applied to simulated data set (WGMG 2003)


Figure 3.4.6
Results for tuning on index with time trend in catchability ( "noized" data set from WGMG 20 (ISVPA, catch-controlled, minimization of MDN - the same options as for NSS herring)


Figure 3.4.7
Pattern of residuals in logarithmic catches in dependence of age diapason used




Figure 3.4.8.
NSS herring. ISVPA. Retrospective runs


Figure 3.4.9.
NSS herring. ISVPA. Bootstrap


Figure 3.5.2.1 Comparison of year-class strength given by the ISVPA and Seastar assessments in 2002 and 2003. Ages refer to 1 Jan 2001.


Figure 3.5.2.2 Comparison of exploitation patterns used by ISVPA and SeaStar in the short-term projection. F values are normalised to 1 as the largest value.


Figure 3.5.2.3 Comparison of year-class strength used by ISVPA and SeaStar in the short-term projection. Ages refer to 1 Jan 2003.

### 4.1 Regulation of the Barents Sea Capelin Fishery

Since 1979, the Barents Sea capelin fishery has been regulated by a bilateral fishery management agreement between Russia (former USSR) and Norway. A TAC has been set separately for the winter fishery and for the autumn fishery. In recent years no autumn fishery has taken place, except for a small Russian experimental fishery. The fishery was closed from 1 May to 15 August until 1984. During the period 1984 to 1986, the fishery was closed from 1 May to 1 September. A minimum landing size of 11 cm has been in force for several years. From the autumn of 1986 to the winter of 1991, and from the autumn 1993 to the winter 1999 no fishery took place. The fishery was re-opened in the winter season 1991 and again in the winter season 1999, on a recovered stock.

In its autumn meeting of 2002, ACFM considered a harvest control rule, which was consistent with the precautionary approach. This rule defined the harvest level based on a maximum probability of $5 \%$ that SSB would fall below $\mathbf{B}_{\text {lim }}$ of 200000 t (corresponding to a catch of 310000 t of pre-spawning capelin in 2003). ACFM also recommended that this harvest control rule be applied in 2003 (see also Section 4.5). During its Autumn 2002 meeting the Mixed Russian Norwegian Fishery Commission decided to set a quota of 310000 t on Barents Sea capelin for the winter season 2002, divided by $60 \%$ (186 000 t) to Norway and $40 \%(124000$ t) to Russia. Moreover, the Commission agreed to adopt a management strategy based on the rule that, with $95 \%$ probability, at least 200000 t of capelin should be allowed to spawn.

### 4.2 Catch Statistics

The international catch by country and season in the years 1965-2002 is given in Table 4.2.1. The catch by age and length groups during the spring season 2002 is given in Table 4.2.2. The total catch in winter 2002 given in Table 4.2.1 was 635000 t . This is 15000 t below the quota set for 2002 . The catch by age and length taken in the Russian experimental fishery during autumn $2002(16000 \mathrm{t})$ is shown in Table 4.2.3.

The final catch statistics for the winter-spring season 2003 are not available. Preliminary figures show that the Norwegian fishery ended around 15 April with about 7000 t of the quota remaining. The Russian fishery ended around 24 April with about 17000 t of the quota remaining.

### 4.3 Stock Size Estimates

### 4.3.1 Larval and 0-group estimates in 2002

Norwegian larval surveys based on Gulf III plankton samples have been carried out in June each year since 1981. The estimated total number of larvae is shown in Table 4.3.1.1. These larval abundance estimates do not show a high correlation with year-class strength at age one, but should reflect the amount of larvae produced each year (Gundersen and Gjøsæter, 1998). The year 1986 was exceptional, in that no larvae were found. This may have been due to late spawning that year, and eggs may have hatched after the survey was carried out. Also in other years some spawning is known to have taken place during the summer, and offspring from such late spawning is not reflected in the larval abundance estimates in Table 4.3.1.1. Since 1997, permission has not been granted to enter the Russian EEZ during the larval survey, and consequently the total larval distribution area has not been covered. The estimate of $22.4 \cdot 10^{12}$ larvae in 2002 is twice as large as the estimate in 2001, at the same level as that obtained in 2000 and far above the average for the period 1981-2002. During the international 0-group surveys in August an area-based index for the abundance of 0group capelin is calculated (Table 4.3.1.1). Gundersen and Gjøsæter (1998) found these indices to be well correlated ( $\mathrm{r}^{2}$ $=0.75$ ) with the 1 -group acoustic estimates for the same year class obtained by the annual capelin acoustic surveys in autumn. Data points up to 1994 were included in this analysis. When this regression is updated with the survey results from 1981-2001 the parameters in the regression were changed slightly and the $r^{2}$ was reduced to 0.68 . Based on this regression, (ln 1 -group estimate $=-1.77+1.19 \cdot \ln 0$-group index), the 0 -group index obtained in 2002 of 327 would correspond to a year-class strength of 170 billion one-year-olds in autumn 2003. A year class of this size would be about $80 \%$ of an average year class in the period 1972-2001.

### 4.3.2 Acoustic stock size estimates in 2002

Two Russian and two Norwegian vessels jointly carried out the 2002 acoustic survey in the period 10 September to 6 October (Bogstad et al, WD.). As previously the Norwegian vessels had restricted access to the Russian EEZ, but since two of the four vessels available to the survey could work in the Russian EEZ, the coverage of the total stock was considered complete. The results from the survey are given in Table 4.3.2.1, and are compared to previous years' results
in Table 4.3.2.2. The stock size was estimated at 2.2 million tonnes. About $60 \%$ ( 1.3 million $t$ ) of the stock biomass consisted of maturing fish (> 14 cm ).

### 4.3.3 Other surveys

During the Norwegian demersal fish survey in February 2002 observations of capelin by acoustics and by pelagic and demersal trawls were made (Røttingen and Gjøsæter, WD). However, no stock size estimate was attempted, due to inadequate coverage and low number of pelagic trawl hauls for identification and sampling purposes. Samples of cod stomachs during this period give valuable information for the modelling of maturing capelin as prey for cod (Bogstad and Gjøsæter, 2001).

### 4.4 Historical stock development

An overview of the development of the Barents Sea capelin stock in the period 1991-2002 is given in Tables 4.4.14.4.7. The methods and assumptions used for constructing the tables are explained in Appendix A to ICES 1995/Assess:9. In that report, the complete time-series back to 1973 can also be found. It should be noted that several of the assumptions and parameter values used in constructing these tables differ from those used in the assessment. For instance, in the assessment model the M-values for immature capelin are calculated using new estimates of the length at maturity and M -values for mature capelin are calculated taking the predation by cod into account. This will also affect the estimates of spawning stock biomass given in the stock summary table (Table 4.4.7). It should be noted that these values, coming from a deterministic model cannot directly be compared to those coming from the probabilistic assessment model (Bifrost) used for this stock. However, as a crude overview of the development of the Barents Sea capelin stock the tables may be adequate.

Estimates of stock in number by age group and total biomass for the period are shown in Table 4.4.1. Catch in numbers-at-age and total landings are shown for the spring and autumn seasons in Tables 4.4.2 and 4.4.3. Natural mortality coefficients by age group for immature and mature capelin are shown in Table 4.4.4. Stock size at 1 January in numbers-at-age and total biomass is shown in Table 4.4.5. Spawning stock biomass per age group is shown in Table 4.4.6. Table 4.4.7 gives an aggregated summary for the entire period 1973-2002.

### 4.5 Stock assessment autumn 2002

As decided by the Northern Pelagic and Blue Whiting Fisheries Working Group at its 2002 meeting (ICES 2002/ACFM:19), the assessment of Barents Sea capelin was left to the parties responsible for the autumn survey, i.e. IMR in Bergen and PINRO in Murmansk, who reported directly to ACFM before its autumn 2002 meeting (Bogstad et al., WD).

A probabilistic projection of the spawning stock to the time of spawning at 1 April 2003 was presented, using the spreadsheet model CapTool (implemented in the @RISK add-on for EXCEL). The projection was based on a probabilistic maturation model with parameters estimated by the model Bifrost (Gjøsæter et al., 2002) with uncertainty taken into account and data on size and composition of the cod stock (from the Arctic Fisheries Working Group, ICES 2002/ACFM:18, but made probabilistic in CapTool in accordance with the risk analysis made by the Arctic Fisheries Working Group).

There is clearly a need for a target biomass reference point for capelin. Calculations of $\mathrm{B}_{\text {target }}$ were attempted, but were not presented because the results were considered preliminary. A $\mathbf{B}_{\mathrm{lim}}\left(\mathrm{SSB}_{\mathrm{lim}}\right)$ management approach was suggested for this stock. As in previous years, the meeting suggested the spawning stock size in 1989 as a $\mathbf{B}_{\text {lim }}$. The rationale behind this was that this biomass produced one of the strongest year classes observed during the period 1972-2001. It should also be noted that this year is within the time range for which quantitative stomach content data are available. It can be argued that the SSB in 1989 was sufficiently large to produce a good year class under favourable recruitment conditions in a "non-herring situation" (Gjøsæter and Bogstad, 1998).

Probabilistic prognoses for the maturing stock from October 1, 2002 until April 1, 2003 were made, with a CV of 0.20 on the abundance estimate. The meeting concluded that capelin recruitment in 2003 would probably not be influenced to any noticeable degree by the stock of young herring now found in the Barents Sea.

ACFM at its autumn 2002 meeting (ICES 2002/CRR:255) took most of the points in the report into account. ACFM agreed to the view that fishing mortality reference points and a $\mathbf{B}_{\mathrm{pa}}$ are not relevant for this stock, and that a target escapement management strategy is the most useful way of ensuring a minimum amount of spawners. Further ACFM
agreed to the strategy adopted of directing the fishery at the spawning stock just prior to spawning, to allow the capelin to be available to predators as long as possible. The idea of a stochastic $\mathbf{B}_{\text {lim }}$ set equal to the modelled density distribution of the spawning stock in 1989 was considered "a good basis for such a reference point in a non-herring situation". Because the assessment method may not yet account for all sources of uncertainty, and there are inconsistencies in the data series ACFM did not adopt the suggested $\mathbf{B}_{\text {lim }}$. Rather, ACFM set a $\mathbf{B}_{\text {lim }}$ of 200000 t , and consequently, advised that a TAC should not exceed 310000 t . This was based on adopting the forecast of the SSB using the limit reference points referred above, and following the harvest control rule that the SSB should fall below $\mathbf{B}_{\mathrm{lim}}$ with a maximum $5 \%$ probability. ACFM also considered that adjustments of the harvest control rule should be further investigated for the purpose of taking better account of the uncertainty in the predicted estimate of spawner abundance, the likely interactions with herring, and the role of capelin as prey.

### 4.6 Management considerations

Since the assessment of the stock is directly based on the acoustic survey conducted annually in September-October, and the main fishing season does not begin until January, advice for this stock must be given during the autumn ACFM meeting and the TAC must be set by the Mixed Norwegian-Russian Fishery Commission during its meeting in November-December. As previously decided by the Northern Pelagic and Blue Whiting Fisheries Working Group, the assessment of Barents Sea capelin is left to the parties responsible for the autumn survey, i.e. IMR in Bergen and PINRO in Murmansk, who will meet in Murmansk after the autumn survey and report directly to the 2003 ACFM autumn meeting.

### 4.7 Sampling

The sampling from scientific surveys and from commercial fishing on capelin in 2002 and winter 2003 is summarised below:

| Investigation | No. of samples | Length measurements | Aged individuals |
| :--- | :---: | :---: | :---: |
| Russian capelin investigation winter 2002 | 18 | 9105 | 900 |
| Russian fishery winter-spring 2002 | 9 | 13145 | 751 |
| Norwegian capelin investigations winter 2002 | 221 | 6578 | 2463 |
| Norwegian fishery winter-spring 2002 | 77 | 7851 | 1700 |
| Acoustic survey autumn 2002 (Norway) | 87 | 6539 | 4830 |
| Acoustic survey autumn 2002 (Russia) | 24 | 22206 | 972 |
| Other samples 2002 (Norway) | 363 | 17491 | 1550 |
| Autumn monitoring 2002 (Russian) | 25 | 37581 | 1250 |
| Russian fishery winter-spring 2003 | 16 | 10349 | 800 |

Table 4.2.1 Barents Sea CAPELIN. International catch (' 000 t ) as used by the Working Group.

| Year | Winter |  |  |  | Summer-Autumn |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Norway | Russia | Others | Total | Norway | Russia | Total |  |
| 1965 | 217 | 7 | 0 | 224 | 0 | 0 | 0 | 224 |
| 1966 | 380 | 9 | 0 | 389 | 0 | 0 | 0 | 389 |
| 1967 | 403 | 6 | 0 | 409 | 0 | 0 | 0 | 409 |
| 1968 | 460 | 15 | 0 | 475 | 62 | 0 | 62 | 537 |
| 1969 | 436 | 1 | 0 | 437 | 243 | 0 | 243 | 680 |
| 1970 | 955 | 8 | 0 | 963 | 346 | 5 | 351 | 1314 |
| 1971 | 1300 | 14 | 0 | 1314 | 71 | 7 | 78 | 1392 |
| 1972 | 1208 | 24 | 0 | 1232 | 347 | 11 | 358 | 1591 |
| 1973 | 1078 | 35 | 0 | 1112 | 213 | 10 | 223 | 1336 |
| 1974 | 749 | 80 | 0 | 829 | 237 | 82 | 319 | 1149 |
| 1975 | 559 | 301 | 43 | 903 | 407 | 129 | 536 | 1439 |
| 1976 | 1252 | 231 | 0 | 1482 | 739 | 366 | 1105 | 2587 |
| 1977 | 1441 | 345 | 2 | 1788 | 722 | 477 | 1199 | 2987 |
| 1978 | 784 | 436 | 25 | 1245 | 360 | 311 | 671 | 1916 |
| 1979 | 539 | 343 | 5 | 887 | 570 | 326 | 896 | 1783 |
| 1980 | 539 | 253 | 9 | 801 | 459 | 388 | 847 | 1648 |
| 1981 | 784 | 428 | 28 | 1240 | 454 | 292 | 746 | 1986 |
| 1982 | 568 | 260 | 5 | 833 | 591 | 336 | 927 | 1760 |
| 1983 | 751 | 374 | 36 | 1161 | 758 | 439 | 1197 | 2358 |
| 1984 | 330 | 257 | 42 | 628 | 481 | 367 | 849 | 1477 |
| 1985 | 340 | 234 | 17 | 590 | 113 | 164 | 278 | 868 |
| 1986 | 72 | 51 | 0 | 123 | 0 | 0 | 0 | 123 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 528 | 156 | 20 | 704 | 31 | 195 | 226 | 929 |
| 1992 | 620 | 247 | 24 | 891 | 73 | 159 | 232 | 1123 |
| 1993 | 402 | 170 | 14 | 586 | 0 | 0 | 0 | 586 |
| 1994 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 |
| 1998 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 |
| 1999 | 46 | 32 | 0 | 78 | 0 | 23 | 23 | 101 |
| 2000 | 283 | 95 | 8 | 386 | 0 | 28 | 28 | 414 |
| 2001 | 368 | 180 | 8 | 557 | 0 | 11 | 11 | 568 |
| 2002 | 391 | 228 | 17 | 635 | 0 | 16 | 16 | 651 |
| $2003{ }^{1}$ | 179 | 107 |  |  |  |  |  |  |

[^2]Table 4.2.2 Barents Sea CAPELIN. International catch in number $\left(10^{6}\right)$ and biomass $(t)$ during the spring season 2002, as used by the Working Group

| Length cm | $\begin{gathered} \text { Age } 1 \\ \mathrm{~N} \end{gathered}$ |  | Age 2 |  | Age 3 |  | Age 4 |  | Age 5+ |  | Sum |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | B | N | B | N | B | N | B | N | B | N | \% | B | \% |
| 5.0-5.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5.5-6.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6.0-6.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6.5-7.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7.0-7.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7.5-8.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8.0-8.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8.5-9.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9.0-9.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9.5-10.0 | 0 | 0 | 27 | 81 | 0 | 0 | 0 | 0 | 0 | 0 | 27 | 0 | 81 | 0 |
| 10.0-10.5 | 0 | 0 | 4 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 13 | 0 |
| 10.5-11.0 | 0 | 0 | 4 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 15 | 0 |
| 11.0-11.5 | 0 | 0 | 4 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 24 | 0 |
| 11.5-12.0 | 0 | 0 | 6 | 172 | 7 | 53 | 0 | 0 | 0 | 0 | 13 | 0 | 224 | 0 |
| 12.0-12.5 | 23 | 0 | 12 | 82 | 24 | 165 | 0 | 0 | 0 | 0 | 59 | 0 | 247 | 0 |
| 12.5-13.0 | 0 | 0 | 0 | 0 | 23 | 220 | 0 | 0 | 0 | 0 | 23 | 0 | 220 | 0 |
| 13.0-13.5 | 0 | 0 | 0 | 1 | 261 | 2662 | 0 | 0 | 0 | 0 | 261 | 1 | 2663 | 0 |
| 13.5-14.0 | 0 | 0 | 11 | 103 | 627 | 7000 | 0 | 0 | 0 | 0 | 638 | 3 | 7103 | 1 |
| 14.0-14.5 | 0 | 0 | 12 | 128 | 609 | 7739 | 30 | 391 | 0 | 0 | 651 | 3 | 8259 | 1 |
| 14.5-15.0 | 0 | 0 | 50 | 593 | 1289 | 18504 | 39 | 588 | 0 | 0 | 1378 | 5 | 19685 | 3 |
| 15.0-15.5 | 0 | 0 | 57 | 756 | 1557 | 25370 | 256 | 4316 | 0 | 0 | 1870 | 7 | 30442 | 5 |
| 15.5-16.0 | 0 | 0 | 45 | 630 | 1381 | 25260 | 685 | 12919 | 0 | 0 | 2112 | 8 | 38809 | 6 |
| 16.0-16.5 | 0 | 0 | 23 | 390 | 1344 | 27926 | 1747 | 38247 | 25 | 489 | 3140 | 12 | 67052 | 11 |
| 16.5-17.0 | 0 | 0 | 68 | 1244 | 1105 | 25621 | 2219 | 53905 | 54 | 1216 | 3446 | 14 | 81985 | 13 |
| 17.0-17.5 | 0 | 0 | 34 | 658 | 756 | 19762 | 2154 | 59203 | 47 | 1143 | 2991 | 12 | 80765 | 13 |
| 17.5-18.0 | 0 | 0 | 0 | 0 | 634 | 18176 | 2425 | 74118 | 61 | 1816 | 3120 | 12 | 94109 | 15 |
| 18.0-18.5 | 0 | 0 | 0 | 0 | 229 | 7679 | 2112 | 72538 | 185 | 5797 | 2526 | 10 | 86013 | 14 |
| 18.5-19.0 | 0 | 0 | 0 | 0 | 100 | 3579 | 1516 | 56606 | 142 | 5023 | 1758 | 7 | 65208 | 10 |
| 19.0-19.5 | 0 | 0 | 0 | 0 | 37 | 1422 | 775 | 31632 | 77 | 3157 | 889 | 4 | 36212 | 6 |
| 19.5-20.0 | 0 | 0 | 0 | 0 | 23 | 1135 | 208 | 8923 | 62 | 2612 | 292 | 1 | 12670 | 2 |
| 20.0-20.5 | 0 | 0 | 0 | 0 | 0 | 0 | 56 | 2613 | 10 | 459 | 66 | 0 | 3072 | 0 |
| 20.5-21.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 21.0-21.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 21.5-22.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 4.2.3 Barents Sea CAPELIN. Russian catch in number $\left(10^{6}\right)$ and biomass $(\mathrm{t})$ during the autumn season 2001, as used by the Working Group

| Length <br> cm | Age 1 |  | Age 2 |  | Age 3 |  | Age 4 |  | Age 5+ |  | Sum |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | N | B | N | B | N | B | N | B | N | B | N | \% | B | \% |
| 5.0-5.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5.5-6.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6.0-6.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6.5-7.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7.0-7.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7.5-8.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8.0-8.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8.5-9.0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 0 |
| 9.0-9.5 | 2 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 7 | 0 |
| 9.5-10.0 | 2 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 7 | 0 |
| 10.0-10.5 | 3 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 11 | 0 |
| 10.5-11.0 | 6 | 26 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 1 | 26 | 0 |
| 11.0-11.5 | 6 | 29 | 2 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 1 | 39 | 0 |
| 11.5-12.0 | 3 | 22 | 9 | 63 | 0 | 0 | 0 | 0 | 0 | 0 | 13 | 1 | 85 | 1 |
| 12.0-12.5 | 0 | 3 | 16 | 117 | 1 | 7 | 0 | 0 | 0 | 0 | 17 | 2 | 126 | 1 |
| 12.5-13.0 | 0 | 0 | 25 | 209 | 0 | 0 | 0 | 0 | 0 | 0 | 25 | 3 | 209 | 1 |
| 13.0-13.5 | 0 | 4 | 38 | 363 | 4 | 43 | 0 | 0 | 0 | 0 | 43 | 5 | 410 | 3 |
| 13.5-14.0 | 0 | 0 | 54 | 587 | 11 | 132 | 0 | 0 | 0 | 0 | 65 | 7 | 719 | 4 |
| 14.0-14.5 | 0 | 0 | 55 | 671 | 25 | 313 | 0 | 0 | 0 | 0 | 80 | 9 | 985 | 6 |
| 14.5-15.0 | 0 | 0 | 39 | 516 | 50 | 712 | 0 | 0 | 0 | 0 | 89 | 10 | 1228 | 8 |
| 15.0-15.5 | 0 | 0 | 26 | 381 | 62 | 997 | 0 | 0 | 0 | 0 | 88 | 10 | 1378 | 9 |
| 15.5-16.0 | 0 | 0 | 11 | 190 | 70 | 1251 | 0 | 0 | 0 | 0 | 81 | 9 | 1441 | 9 |
| 16.0-16.5 | 0 | 0 | 7 | 141 | 72 | 1426 | 1 | 16 | 0 | 0 | 80 | 9 | 1583 | 10 |
| 16.5-17.0 | 0 | 0 | 7 | 143 | 56 | 1239 | 1 | 18 | 0 | 0 | 63 | 7 | 1401 | 9 |
| 17.0-17.5 | 0 | 0 | 1 | 19 | 61 | 1546 | 1 | 19 | 0 | 0 | 63 | 7 | 1583 | 10 |
| 17.5-18.0 | 0 | 0 | 0 | 12 | 49 | 1377 | 0 | 0 | 0 | 0 | 50 | 6 | 1389 | 9 |
| 18.0-18.5 | 0 | 0 | 2 | 55 | 37 | 1171 | 2 | 47 | 0 | 0 | 40 | 5 | 1273 | 8 |
| 18.5-19.0 | 0 | 0 | 1 | 44 | 29 | 998 | 1 | 24 | 0 | 0 | 31 | 4 | 1067 | 7 |
| 19.0-19.5 | 0 | 0 | 0 | 0 | 16 | 608 | 2 | 51 | 0 | 0 | 18 | 2 | 659 | 4 |
| 19.5-20.0 | 0 | 0 | 0 | 0 | 5 | 193 | 1 | 31 | 0 | 0 | 6 | 1 | 225 | 1 |
| 20.0-20.5 | 0 | 0 | 0 | 0 | 2 | 110 | 1 | 37 | 0 | 0 | 3 | 0 | 147 | 1 |
| 20.5-21.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 21.0-21.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 21.5-22.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 4.3.1.1 Barents Sea CAPELIN. Larval abundance estimate $\left(10^{12}\right)$ in June, and 0 -group index in August.

| Larval <br> Year Abundance |  | $\begin{gathered} 0 \text {-group } \\ \text { index } \end{gathered}$ |
| :---: | :---: | :---: |
| 1981 | 9.7 | 570 |
| 1982 | 9.9 | 393 |
| 1983 | 9.9 | 589 |
| 1984 | 8.2 | 320 |
| 1985 | 8.6 | 110 |
| 1986 | 0.0 | 125 |
| 1987 | 0.3 | 55 |
| 1988 | 0.3 | 187 |
| 1989 | 7.3 | 1300 |
| 1990 | 13.0 | 324 |
| 1991 | 3.0 | 241 |
| 1992 | 7.3 | 26 |
| 1993 | 3.3 | 43 |
| 1994 | 0.1 | 58 |
| 1995 | 0.0 | 43 |
| 1996 | 2.4 | 291 |
| 1997 | 6.9 | 522 |
| 1998 | 14.1 | 428 |
| 1999 | 36.5 | 722 |
| 2000 | 19.1 | 303 |
| 2001 | 10.7 | 221 |
| 2002 | 22.4 | 327 |

Table 4.3.2.1 Barents Sea CAPELIN. Estimated stock size from the acoustic survey in September-October 2002. Based on TS value $19.1 \log \mathrm{~L}-74.0 \mathrm{~dB}$, corresponding to $\sigma=5.0 \cdot 10^{7} \cdot \mathrm{~L}^{1.91}$.

| Age/Year class |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Length (cm) | 1 | 2 | 3 |  |  | Biomass |  |
|  | 2001 | 2000 | 1999 | 1998 | $\left(10^{6}\right)$ | $\left(10^{3} \mathrm{t}\right)$ |  |
| $6.5-7.0$ | 2 |  |  |  | 2 | 0.0 | 1.0 |
| 7.0-7.5 | 214 |  |  |  | 214 | 0.3 | 1.6 |
| 7.5-8.0 | 1674 |  |  |  | 1674 | 2.6 | 1.6 |
| $8.0-8.5$ | 8844 |  |  |  | 8844 | 18.1 | 2.0 |
| 8.5 - 9.0 | 7407 |  |  |  | 7407 | 18.8 | 2.5 |
| 9.0-9.5 | 6206 |  |  |  | 6206 | 19.6 | 3.2 |
| 9.5-10.0 | 4797 |  |  |  | 4797 | 16.7 | 3.5 |
| 10.0-10.5 | 7130 | 55 |  |  | 7185 | 30.5 | 4.2 |
| 10.5-11.0 | 10471 | 758 |  |  | 11229 | 53.5 | 4.8 |
| 11.0-11.5 | 8102 | 2594 |  |  | 10697 | 58.4 | 5.5 |
| 11.5-12.0 | 4033 | 9706 |  |  | 13738 | 87.5 | 6.4 |
| 12.0-12.5 | 451 | 17173 | 61 |  | 17684 | 138.0 | 7.8 |
| 12.5-13.0 | 26 | 17854 | 286 |  | 18166 | 162.5 | 8.9 |
| 13.0-13.5 | 204 | 14255 | 1486 |  | 15945 | 158.0 | 9.9 |
| 13.5-14.0 | 78 | 11089 | 2076 |  | 13243 | 154.4 | 11.7 |
| 14.0-14.5 | 64 | 8470 | 5812 |  | 14346 | 190.0 | 13.2 |
| 14.5-15.0 |  | 3824 | 6503 |  | 10327 | 155.3 | 15.0 |
| 15.0-15.5 |  | 2578 | 6422 |  | 9000 | 155.2 | 17.2 |
| 15.5-16.0 |  | 916 | 6284 | 9 | 7209 | 142.8 | 19.8 |
| 16.0-16.5 |  | 392 | 6248 | 48 | 6688 | 152.6 | 22.8 |
| 16.5-17.0 |  | 424 | 5274 | 32 | 5730 | 146.3 | 25.5 |
| 17.0-17.5 |  | 292 | 4007 | 47 | 4346 | 122.8 | 28.3 |
| 17.5-18.0 |  | 355 | 3193 | 91 | 3639 | 115.7 | 31.8 |
| 18.0-18.5 |  | 61 | 1735 | 86 | 1882 | 67.5 | 35.8 |
| 18.5-19.0 |  |  | 642 | 134 | 776 | 29.8 | 38.3 |
| 19.0-19.5 |  |  | 181 | 58 | 239 | 10.3 | 43.0 |
| 19.5-20.0 |  |  |  | 60 | 60 | 2.6 | 43.0 |
| 20.0-20.5 |  |  |  | 8 | 8 | 0.4 | 51.1 |
| 20.5-21.0 |  |  |  | 2 | 2 | 0.1 | 55.8 |
| TSN (10 ${ }^{6}$ ) | 59704 | 90794 | 50211 | 577 | 201286 |  |  |
| TSB ( $10^{3} \mathrm{t}$ ) | 234.3 | 918.6 | 1037.1 | 20.2 |  | 2210.2 |  |
| Mean length (cm) | 9.90 | 13.10 | 15.80 | 18.20 | 12.80 |  |  |
| Mean weight (g) | 3.9 | 10.1 | 20.7 | 35.0 |  |  | 11.0 |
| $\operatorname{SSN}\left(10^{6}\right)$ | 64 | 17312 | 46301 | 575 | 64252 |  |  |
| SSB $\left(10^{3} \mathrm{t}\right)$ | 0.8 | 273.1 | 995.5 | 20.1 |  | 1291.4 |  |
|  |  | d on TS | lue: 19.1 | L-74.0 | correspon | ing to $\sigma=5$. | $\cdot 10^{-7} \cdot \mathrm{~L}^{1.9}$ |

Table 4.3.2.2 Barents Sea CAPELIN. Stock size in numbers by age, total stock biomass and biomass of the maturing component. Stock in numbers (unit: $10^{9}$ ) and stock and maturing stock biomass (unit: $10^{3}$ tonnes) are given at 1 October.

| Year |  | Stock in numbers $\left(10^{9}\right)$ |  |  |  |  |  |  |  |  | Stock in weight |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | $(000$ t $)$ |  |  |  |  |  |  |  |
|  | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 | Total | Total | Maturing |  |  |  |  |  |
| 1973 | 528 | 375 | 40 | 17 | 0 | 961 | 5144 | 1350 |  |  |  |  |  |
| 1974 | 305 | 547 | 173 | 3 | 0 | 1029 | 5733 | 907 |  |  |  |  |  |
| 1975 | 190 | 348 | 296 | 86 | 0 | 921 | 7806 | 2916 |  |  |  |  |  |
| 1976 | 211 | 233 | 163 | 77 | 12 | 696 | 6417 | 3200 |  |  |  |  |  |
| 1977 | 360 | 175 | 99 | 40 | 7 | 681 | 4796 | 2676 |  |  |  |  |  |
| 1978 | 84 | 392 | 76 | 9 | 1 | 561 | 4247 | 1402 |  |  |  |  |  |
| 1979 | 12 | 333 | 114 | 5 | 0 | 464 | 4162 | 1227 |  |  |  |  |  |
| 1980 | 270 | 196 | 155 | 33 | 0 | 654 | 6715 | 3913 |  |  |  |  |  |
| 1981 | 403 | 195 | 48 | 14 | 0 | 660 | 3895 | 1551 |  |  |  |  |  |
| 1982 | 528 | 148 | 57 | 2 | 0 | 735 | 3779 | 1591 |  |  |  |  |  |
| 1983 | 515 | 200 | 38 | 0 | 0 | 754 | 4230 | 1329 |  |  |  |  |  |
| 1984 | 155 | 187 | 48 | 3 | 0 | 393 | 2964 | 1208 |  |  |  |  |  |
| 1985 | 39 | 48 | 21 | 1 | 0 | 109 | 860 | 285 |  |  |  |  |  |
| 1986 | 6 | 5 | 3 | 0 | 0 | 14 | 120 | 65 |  |  |  |  |  |
| 1987 | 38 | 2 | 0 | 0 | 0 | 39 | 101 | 17 |  |  |  |  |  |
| 1988 | 21 | 29 | 0 | 0 | 0 | 50 | 428 | 200 |  |  |  |  |  |
| 1989 | 189 | 18 | 3 | 0 | 0 | 209 | 864 | 175 |  |  |  |  |  |
| 1990 | 700 | 178 | 16 | 0 | 0 | 894 | 5831 | 2617 |  |  |  |  |  |
| 1991 | 402 | 580 | 33 | 1 | 0 | 1016 | 7287 | 2248 |  |  |  |  |  |
| 1992 | 351 | 196 | 129 | 1 | 0 | 678 | 5150 | 2228 |  |  |  |  |  |
| 1993 | 2 | 53 | 17 | 2 | 2 | 75 | 796 | 330 |  |  |  |  |  |
| 1994 | 20 | 3 | 4 | 0 | 0 | 28 | 200 | 94 |  |  |  |  |  |
| 1995 | 7 | 8 | 2 | 0 | 0 | 17 | 193 | 118 |  |  |  |  |  |
| 1996 | 82 | 12 | 2 | 0 | 0 | 96 | 503 | 248 |  |  |  |  |  |
| 1997 | 99 | 39 | 2 | 0 | 0 | 140 | 911 | 312 |  |  |  |  |  |
| 1998 | 179 | 73 | 11 | 1 | 0 | 263 | 2056 | 931 |  |  |  |  |  |
| 1999 | 156 | 101 | 27 | 1 | 0 | 285 | 2776 | 1718 |  |  |  |  |  |
| 2000 | 449 | 111 | 34 | 1 | 0 | 595 | 4273 | 2099 |  |  |  |  |  |
| 2001 | 114 | 219 | 31 | 1 | 0 | 364 | 3630 | 2019 |  |  |  |  |  |
| 2002 | 60 | 91 | 50 | 1 | 0 | 201 | 2210 | 1290 |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 4.4.1 Barents Sea CAPELIN. Estimated stock size in numbers (unit: $10^{9}$ ) by age group and total, and biomass (' 000 t ) of total stock, by 1 August, back-calculated from the survey in September-October.

| Age |  | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 | 3.1 | 29.5 | 8.3 | 88.9 | 111.8 | 188.4 | 171.4 | 474.7 | 128.0 |
|  | 2 | 73.0 | 5.1 | 9.4 | 12.5 | 44.2 | 76.5 | 111.5 | 116.8 | 246.6 |
|  | 3 | 25.3 | 6.4 | 1.6 | 2.2 | 2.2 | 12.1 | 27.9 | 35.9 | 33.0 |
|  | 4 | 3.7 | 0.3 | 0.4 | 0.1 | 0.1 | 0.7 | 0.9 | 0.8 | 1.2 |
|  | 5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.1 | 0.1 |
| Sum |  | 105.0 | 41.4 | 19.7 | 103.7 | 158.3 | 277.8 | 311.7 | 628.4 | 408.8 |
| Biomass | 991 | 259 | 189 | 467 | 866 | 1860 | 2580 | 3840 | 3480 | 224.7 |

Table 4.4.2 Barents Sea CAPELIN. Catch in numbers (unit: $10^{9}$ ) by age group and total landings (' 000 t ) in the spring season.

| Age |  | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 |
|  | 2 | 0.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.5 |
|  | 3 | 4.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 1.6 | 5.5 | 7.6 |
|  | 4 | 26.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 1.2 | 8.4 | 12.1 |
|  | 5 | 1.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 1.0 | 2.2 |
| Sum | 33.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 3.0 | 15.1 | 22.5 | 25.3 |
| Landings | 586 | 0 | 0 | 0 | 0 | 0 | 78 | 386 | 557 | 635 |

Table 4.4.3 Barents Sea CAPELIN. Catch in numbers (unit: $10^{9}$ ) by age group and total landings (' 000 t ) in the autumn season.

| Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.4 | 0.1 | 0.0 |
|  | 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.9 | 0.9 | 0.4 |
|  | 3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.4 | 0.2 |
|  |  |  |  |  |  |  |  |  |  |  |
|  | 4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
|  | 5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Sum | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 1.6 | 1.5 | 0.6 | 0.0 |
| Landings | 0 | 0 | 0 | 0 | 1 | 1 | 23 | 28 | 11 | 16 |

Table 4.4.4 Barents Sea CAPELIN. Natural mortality coefficients (per month) for immature fish $\left(\mathrm{M}_{\mathrm{imm}}\right)$, used for the whole year, and for mature fish (per season) ( $\mathrm{M}_{\text {mat }}$ ) used January to March, by age group and average for age groups 1-5.


Table 4.4.4 (Continued)

| Age | 1998 |  |  | 1999 |  | 2000 |  | 2001 |  | 2002 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{M}_{\text {imm }}$ | $\mathrm{M}_{\text {mat }}$ | $\mathrm{M}_{\text {imm }}$ | $\mathrm{M}_{\text {mat }}$ |  |  |  |  |  |  |  |
|  | 1 | 0.026 | 0.077 | 0.047 | 0.142 | 0.028 | 0.083 | 0.060 | 0.180 | 0.060 | 0.180 |
|  | 2 | 0.026 | 0.077 | 0.047 | 0.142 | 0.028 | 0.083 | 0.060 | 0.180 | 0.060 | 0.180 |
|  | 3 | 0.071 | 0.212 | 0.025 | 0.074 | 0.026 | 0.079 | 0.040 | 0.120 | 0.040 | 0.120 |
|  | 4 | 0.071 | 0.212 | 0.025 | 0.074 | 0.026 | 0.079 | 0.040 | 0.120 | 0.040 | 0.120 |
|  | 5 | 0.071 | 0.212 | 0.025 | 0.074 | 0.026 | 0.079 | 0.040 | 0.120 | 0.040 | 0.120 |
| Avr |  | 0.053 | 0.158 | 0.034 | 0.101 | 0.027 | 0.080 | 0.048 | 0.144 | 0.048 | 0.144 |

Table 4.4.5 Barents Sea CAPELIN. Estimated stock size in numbers (unit: $10^{9}$ ) by age group and total, and biomass (' 000 t ) of total stock, by 1 January.

| Age |  | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 | 9.2 | 120.3 | 13.8 | 118.2 | 172.0 | 225.5 | 238.5 | 576.1 | 194.7 | 102.3 |
|  | 2 | 293.7 | 1.4 | 10.8 | 5.7 | 72.5 | 82.2 | 165.8 | 135.3 | 413.3 | 94.6 |
|  | 3 | 162.6 | 33.3 | 1.9 | 6.5 | 10.2 | 32.5 | 67.3 | 88.1 | 100.9 | 182.6 |
|  | 4 | 89.2 | 9.8 | 2.4 | 1.4 | 1.8 | 1.6 | 8.5 | 24.7 | 31.1 | 27.0 |
|  | 5 | 0.5 | 1.3 | 0.1 | 0.3 | 0.1 | 0.1 | 0.5 | 0.8 | 0.7 | 0.9 |
| Sum | 555.2 | 166.1 | 28.9 | 132.2 | 256.6 | 341.9 | 480.6 | 824.9 | 740.6 | 407.5 |  |
| Biomass | 4372 | 737 | 156 | 313 | 779 | 1240 | 2456 | 3571 | 4558 | 3539 |  |

Table 4.4.6 Barents Sea CAPELIN. Estimated spawning stock biomass ('000 t) by 1 April.

| Age |  | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 2 | 0 | 0 | 1 | 3 | 1 | 1 | 2 | 24 | 5 | 0 |
|  | 3 | 129 | 34 | 15 | 71 | 175 | 217 | 650 | 819 | 943 | 867 |
|  | 4 | 331 | 60 | 38 | 24 | 49 | 34 | 193 | 472 | 539 | 339 |
|  | 5 | 0 | 11 | 1 | 7 | 2 | 2 | 10 | 0 | 0 | 9 |
| Sum |  | 460 | 105 | 55 | 105 | 228 | 254 | 856 | 1315 | 1487 | 1215 |

Table 4.4.7 Barents Sea CAPELIN. Stock summary table. Recruitment (number of 1-year-old fish (unit: $10^{9}$ ) and stock biomass ( ${ }^{\circ} 000 \mathrm{t}$ ) given at 1 August, spawning stock ( ${ }^{\circ} 000 \mathrm{t}$ ) at time of spawning ( 1 April). Landings (' 000 t ) are the sum of the total landings in the two fishing seasons within the year indicated. The SSB is obtained by projecting the stock forward assuming a natural mortality that does not take the current predation mortality fully into account.

| Year | Stock biomass | $\begin{gathered} \text { Recruit- } \\ \text { ment Age } 1 \\ \hline \end{gathered}$ | Spawning stock biomass | Landings |
| :---: | :---: | :---: | :---: | :---: |
| 1965 |  |  |  | 224 |
| 1966 |  |  |  | 389 |
| 1967 |  |  |  | 409 |
| 1968 |  |  |  | 537 |
| 1969 |  |  |  | 680 |
| 1970 |  |  |  | 1314 |
| 1971 |  |  |  | 1392 |
| 1972 | 5831 |  |  | 1592 |
| 1973 | 6630 | 1140 | 1242 | 1336 |
| 1974 | 7121 | 737 | 343 | 1149 |
| 1975 | 8841 | 494 | 90 | 1439 |
| 1976 | 7584 | 433 | 1147 | 2587 |
| 1977 | 6254 | 830 | 890 | 2987 |
| 1978 | 6119 | 855 | 460 | 1916 |
| 1979 | 6576 | 551 | 193 | 1783 |
| 1980 | 8219 | 592 | 87 | 1648 |
| 1981 | 4489 | 466 | 1731 | 1986 |
| 1982 | 4205 | 611 | 546 | 1760 |
| 1983 | 4772 | 612 | 47 | 2358 |
| 1984 | 3303 | 183 | 171 | 1477 |
| 1985 | 1087 | 47 | 106 | 868 |
| 1986 | 157 | 9 | 13 | 123 |
| 1987 | 107 | 46 | 16 | 0 |
| 1988 | 361 | 22 | 11 | 0 |
| 1989 | 771 | 195 | 141 | 0 |
| 1990 | 4901 | 708 | 179 | 0 |
| 1991 | 6647 | 415 | 1584 | 929 |
| 1992 | 5371 | 396 | 998 | 1123 |
| 1993 | 991 | 3 | 460 | 586 |
| 1994 | 259 | 30 | 105 | 0 |
| 1995 | 189 | 8 | 55 | 0 |
| 1996 | 467 | 89 | 105 | 0 |
| 1997 | 866 | 112 | 228 | 1 |
| 1998 | 1860 | 188 | 254 | 1 |
| 1999 | 2580 | 171 | 856 | 106 |
| 2000 | 3840 | 475 | 1315 | 414 |
| 2001 | 3480 | 128 | 1487 | 568 |
| 2002 | 2122 | 67 | 1215 | 651 |

## $5.1 \quad$ The Fishery

### 5.1.1 Regulation of the fishery

The fishery depends upon maturing capelin, i.e. that part of each year class which spawns at age 3 as well as those fish at age 4 that did not mature and spawn at age 3 . The abundance of the immature component is difficult to assess before their recruitment to the adult stock at ages 2 and 3 . This is especially true of the age 3 immatures.

The fishery of the Iceland-East Greenland-Jan Mayen capelin has been regulated by preliminary catch quotas set prior to each fishing season (July-March). Predictions of TACs have been computed based on estimates from surveys of the abundance of 1- and 2-year-old capelin, carried out in the autumn of the previous year. The process includes historical relationships between such data and the back-calculated abundance of the same year classes, growth rate and stock in numbers, natural mortality, and the provision of a remaining spawning stock biomass of 400000 t . Final catch quotas for each season have then been set according to the results of acoustic surveys on the maturing, fishable stock, carried out in autumn (October-November) and/or winter (January/February) in that fishing season. A more detailed description of the method is given in Section 1.3.5. A summary of the results of this catch regulation procedure is given in Table 5.1.1.1

Over the years, fishing has not been permitted during April to late June and the season opened in July/August or later, depending on the state of the stock. Due to very low stock abundance there was a fishing ban lasting from December 1981 to November 1983. In addition, areas with high abundances of juvenile age 1 and 2 capelin (in the shelf region off NW-, N - and NE-Iceland) have usually been closed to the summer and autumn fishery.

### 5.1.2 The fishery in the $2002 / 2003$ season

In accordance with a previously determined procedure, ACFM recommended that the preliminary TAC should not exceed 700000 t . This is $2 / 3$ of the total TAC predicted for the $2002 / 2003$ season, i.e. 1040000 t . This advice was accepted by all parties concerned.

The season opened on 20 June and the fishery began in deep waters north of the shelf edge, northeast and north of Iceland. This time the fishing grounds soon shifted to the northwest and north. Catch rates were low in the beginning but improved in July as the capelin migrated north. Towards the end of July the northward migration stopped, the capelin began moving back south again, and soon scattered. By the end of July, the total catch was 290000 t , mainly taken in that month on both sides of the Iceland/Greenland EEZ between about $68^{\circ} \mathrm{N}$ and $69^{\circ} 30^{\prime} \mathrm{N}$. After July the capelin remained scattered and no catches were made for the rest of the year, except for $50000 t$ taken just off the shelf break off the eastern north coast of Iceland in December.

The total catch in the 2002 summer and autumn season was approximately 340000 t .
In January 2003, large fishable concentrations of adult capelin were located in deep waters off the shelf northeast of Iceland and resulted in a successful fishery. During January and the first half of February, parts of these capelin concentrations were located much farther offshore and occupied a much larger area than usual. Their southward speed of migration was unusually slow. Thus, by mid-February no capelin had been found south of $65^{\circ} 30^{\prime} \mathrm{N}$ which is about 60 nautical miles further north than usual. This deviation from the 'normal' distribution and migration pattern seems to be linked with abnormally high temperatures north and east of Iceland. The first spawning migration arrived in the shallow coastal waters off SE-Iceland during the last week of February and then migrated rapidly west along the coast to spawn west of Iceland., Although much of the capelin, which arrived later, also continued west along the coast, capelin spawned along the entire south coast as well as west of Iceland in March and early April 2003 .

A total catch of approximately 450000 t of capelin was taken in deep waters east of Iceland, before the first spawning migration approached the shallow waters off the eastern south coast. As usual, catch rates were high in the Icelandic coastal area and by mid-March most of the TAC of 1000000 t , set for the 2002/2003 season, had been taken.

In late March and early April additional spawning migrations appeared on the banks west of the Vestfirdir Peninsula (NW-Iceland). The size of these migrations was not assessed and they were not fished.

The total catch during the 2003 winter season was 648000 t .

The total annual catch of capelin in the Iceland-East Greenland-Jan Mayen area since 1964 is given by weight, season, and fleet in Table 5.2.1.

The total catch in numbers during the summer/autumn 1981-2002 and winter 1982-2003 seasons is given by age and years in Tables 5.2.2 and 5.2.3.

The distribution of the catch during the summer-autumn 2002 and winter 2003 seasons is given by length groups at age in Tables 5.2.4 and 5.2.5.

### 5.3 Surveys of Stock Abundance

### 5.3.1 0-group surveys

The distribution and abundance of 0-group capelin in the Iceland-East Greenland-Jan Mayen area have been recorded during surveys carried out in August since 1970. The survey methods and computations of abundance indices were described by Vilhjálmsson and Fridgeirsson (1976). The abundance indices of 0 -group capelin, divided according to areas, are given in Table 5.3.1.1.

Acoustic estimates of the abundance of 1 -group capelin were made during the August 0 -group surveys (e.g. Vilhjálmsson 1994). However, due to the large variability in this time-series, the August survey data have not been used for stock projections for more than a decade. Directed collection of data on juvenile ages 1 and 2 capelin in August has been discontinued. The abundance of 1-group capelin by number, mean length and weight for 1983-2001 is given in Table 5.3.1.2.

### 5.3.2 Stock abundance in autumn 2002 and winter 2003

### 5.3.2.1 The adult fishable stock

An attempt to estimate the adult fishable stock was not made in the autumn of 2002. An acoustic survey was carried out in the latter half of November in order to assess the abundance of immature age 1 and 2 capelin (year classes 2000 and 2001). Although the survey recorded juvenile capelin in their normal distribution area over the outer shelf northwest and north of the Vestfirdir Peninsula and the western north coast of Iceland, capelin concentrations were extremely scattered and the total estimate was very low ( $<15^{9}$ individuals, mostly of the 2001 year class). Similar to autumn 2002, few adult capelin were found (WD by H.Vilhjálmsson).

During 7-22 January 2002, the abundance of mature capelin was assessed in deep waters northeast of Iceland and off the northern east coast. With regard to weather, surveying conditions were reasonably good in January 2002. Capelin occupied an unusually wide distribution, with parts of the stock located up to 70 nautical miles off the shelf break. Furthermore, the stock was distributed unevenly with dense and, in places, large concentrations alternating with fairly wide areas of low density.

It is common knowledge that capelin migrate south off the east coast of Iceland in January. However, these migrations do not maintain a steady pace and may even stop altogether for a day or two. Consequently, the speed of migration has to be considered when assessing stock abundance during the survey period. Although capelin were migrating south during the survey, the position and progress of the fishery indicated that the speed of migration was very slow. Thus, no capelin were located south of about $65^{\circ} 30^{\prime} \mathrm{N}$ by the end of the survey in the third week of January.

Although three runs were made within the distribution area, none of these covered the entire area. This was caused mainly by occasional breaks caused by stormy weather. After a thorough scrutiny of the entire dataset, it was concluded that an average of all three surveys represented the best estimate of the total capelin biomass.

The total biomass of adult capelin east of Iceland in January 2003 was estimated to be 760000 t . With the exception of small numbers of age group 2 (year class 2001) in a limited area off the eastern north coast of Iceland, practically no immature capelin were recorded in January 2003. Details of the January 2003 acoustic estimates of adult capelin are given in Table 5.3.2.1.1 (WD by H. Vilhjalmsson)

Because of the uncertainties associated with the January estimate, as well as the very high January catch rates compared
to previous years, it was decided to run another survey when the first spawning migrations were beginning to enter the near-shore area off the eastern south coast in February. Because juveniles of the 2001 and 2000 year classes were not found, it was decided to conduct another survey for juveniles after the adult capelin assessment.

The second survey was carried out in 17 February-13 March 2003. On arrival in the area east of Iceland it was observed that the southernmost border of the spawning migrations east of Iceland had only shifted south by 20-30 nautical miles and were still north of $65^{\circ} \mathrm{N}$. During the last week of February, the capelin slowly migrated to the southwest towards the shelf break east of Iceland. The first migration did not arrive at the southeast coast until the end of February.

The first migration quickly moved very close inshore and migrated rapidly west along the coast. Consequently it was difficult to assess because some of the capelin were too close to land for safe navigation and some fish were out of echosounder range due to their near-surface distribution. Although the assessment of this part of the spawning stock was repeated six times, it involved making several unsupported assumptions. The accepted result was slightly higher than 180000 t and represents about $35 \%$ of the total abundance of the spawning stock.

The remaining migrations were assessed in deeper waters and further from the shore off SE-Iceland, as well as just off the shelf break off the east coast, during 4-10 March under good surveying conditions. The total acoustic estimate was 528000 t , which is about 100000 t greater than that obtained during the January survey (WD by H. Vilhjálmsson).

Details of the February-March 2003 acoustic estimates of adult capelin are given in Table 5.3.2.1.2.

### 5.3.2.2 Estimates of immature capelin

An acoustic survey was carried out in the latter half of November 2002 to assess the abundance of immature ages 1 and 2 capelin (year classes 2000 and 2001). Although the survey recorded juvenile capelin in their normal distribution areas over the outer shelf northwest and north of the Vestfirdir Peninsula and the western north coast of Iceland, capelin concentrations were extremely scattered and the total estimate was very low ( $<15^{9}$ individuals, mostly of the 2001 year class). As in autumn 2002, hardly any adult capelin were located.

The January 2003 suvey of the mature fishable stock also recorded immature capelin, mainly of the 2001 year class, near the shelf edge off the eastern north coast of Iceland between about $18^{\circ} \mathrm{W}$ and $15^{\circ} \mathrm{W}$. Some immature fish of the 2000 year class were also recorded near the shelf break further south. However, their numbers were very low.

After the completion of the adult fishable stock survey off SE- and E-Iceland in early March 2003, the survey continued further north east of Iceland and then west off the north coast. Dense concentrations of immature 3-group capelin of the 2000 year class were recorded at the edge of the shelf off the northern east coast. However, these capelin were in a very narrow area and their total abundance was $3.5 * 10^{3}$ individuals.

No juvenile capelin were recorded off the western north coast and off the Vestfirdir Peninsula. However, having crossed the deepest part of the Denmark Strait, the survey recorded scattered concentrations of capelin along the western slope, near the Iceland/Greenland EEZ division line. Samples of these recordings consisted purely of age 2 capelin of the 2001 year class. Drift ice and stormy weather precluded continuing and the survey was terminated in the third week of March.

A survey, lasting only one week due to a shortage of ship time, was conducted in the area west of Vestfirdir, between $66^{\circ} \mathrm{N}$ and $68^{\circ} \mathrm{N}$ and reaching west to $30-31^{\circ} \mathrm{W}$ during 12-18 April 2003. Considerable and in some places very dense capelin concentrations were recorded during this survey, mostly on either side of the deep Iceland-Greenland Channel, around $66^{\circ} \mathrm{N}$ where very few capelin were seen two weeks earlier. West of there capelin were more scattered and difficult to separate from echoes of other organisms. These recordings consisted exclusively of age 2 s and translated into $61.5 * 10^{9}$ individuals. The result of this assessment is given in Table 5.3.2.2.1. Estimates of immature 2- and 3group capelin recorded earlier off NE- and E-Iceland, have been added (WD by H. Vilhjálmsson).

### 5.4 Historical Stock Abundance

Historical estimates of stock abundance are based on the "best" acoustic estimates of the abundance of maturing capelin in autumn and/or winter surveys. The "best" in each case is defined as that estimate on which the final TAC was based. Taking account of the catch in numbers and a monthly natural mortality rate of $\mathrm{M}=0.035$ (ICES 1991/Assess:17), abundance estimates of each age group are then projected to the appropriate point in time. Since natural mortality rates of juvenile capelin are unknown, their abundance by number has been projected using the same natural mortality rate.

The annual abundance by number and weight-at-age for mature and immature capelin in the Iceland-East GreenlandJan Mayen area has been calculated with reference to 1 August and 1 January of the following year for the 1978/792002/03 seasons. The results are given in Tables 5.4.1 and 5.4.2 (1 August and 1 January, respectively). Table 5.4.2 also gives the remaining spawning stock by number and biomass in March/April 1979-2003.

The observed annual mean weight-at-age, obtained from catch and survey data from January, was used to calculate the stock biomass on 1 January. The observed average weight-at-age of adult capelin in autumn (Table 5.5.1.2) is used to calculate mature stock biomass in summer. Because there is a small weight increase among mature capelin in February and March, the remaining spawning stock biomass is underestimated.

### 5.5 Stock Prognoses

### 5.5.1 Stock prognosis and TAC in the 2002/2003 season

The models (ICES 1993/Assess:6; Section 3.1.5) for predicting the numbers of maturing capelin of ages 2 and 3 from the November 2001 acoustic assessment of the 1999 and 1998 year classes gave estimates of 77.2 and $17.3 * 10^{9}$ maturing 2- and 3-group capelin on 1 August 2002. The data used in the comparison between abundance of age groups are given in Table 5.5.1.1, and the mean weights of maturing capelin in autumn in Table 5.5.1.2.

During the last ten years the weight-at-age of adult capelin has been inversely related to adult stock abundance. Simple linear regressions of these data result in $\mathrm{r}^{2}=0.64$ and 0.68 for age groups 2 and 3 , respectively. The two regression plots are shown in Figure 5.5.2.2. Applying the appropriate regression equations, $y=-0.034 x+19.4 ; r^{2}=0.64 ; p<0.05$ for the younger component, and $y=-0.068 x+29.0 ; r^{2}=0.68 ; p<0.05$ for the older one and using the predicted abundance of age groups 2 and 3 on 1 August 2002 combined, i.e. $95.6 * 10^{9}$ fish, results in estimated mean weights of 16.1 and 22.4 g for age groups 2 and 3 , respectively.

The fishable stock biomass, obtained by multiplying the stock in numbers by the predicted mean weight of maturing capelin in autumn, was projected forward to spawning time in March 2003 assuming a monthly $\mathrm{M}=0.035$ and a remaining spawning stock of 400000 t . This resulted in a predicted TAC of 1050000 t spread evenly over August 2002-March 2003 (Table 5.5.1.3). Using the same approach as in previous years, i.e. that the preliminary TAC be set at $2 / 3$ of the predicted total for the season, the WG recommended that a preliminary TAC for the 2002/2003 capelin fishery be set at 690000 t .

According to the January-March 2003 survey results described in Section 5.3.2, the fishable spawning stock was estimated as $28.3 * 10^{9}$ fish on 1 March 2003. At that time the observed mean weight in the fishable stock was 18.6 g and the stock biomass was 530000 t (Table 5.3.2.1.2). With the usual prerequisites of a monthly natural mortality rate of 0.035 and a remaining spawning stock of 400000 t , the above abundance estimate indicated that a TAC of 125000 t was available for the remainder of the 2003 winter fishery. Adding this to the catch of 875000 t already taken from June 2002-February 2003 resulted in a total TAC of 1000000 t for the 2002/2003 season. In practical terms, this TAC is equal to that predicted for the $2002 / 2003$ season in terms of weight although considerably lower by number, especially for the older age group. The similarity in the biomass prediction is explained by better growth and higher mean weight of the 2000 year class.

About 10000 t of the TAC remained at the end of the winter fishery. As a result, 410000 t of capelin spawned in 2003.

### 5.5.2 Stock prognosis and assessment for the 2003/2004 season

Calculations of expected TAC for the 2003/2004 season, based on the method described in Section 3.1.5 of ICES 1993/Assess: 6 and data from Table 5.5.1.1, were used to predict the abundance of maturing capelin of ages 2 and 3 on 1 August 2003.

An updated linear regression of the measured abundance of 1-group capelin $\left(\mathrm{N}_{1}\right)$ on the back-calculated abundance of mature 2-group fish ( $N_{2 \text { mat }}$ ) gives $y=0.570 x+20.6 ; R^{2}=0.84 ; p<0.05$. Similarly for the older stock component, where $\mathrm{N}_{2 \text { tot }}$ is regressed on $\mathrm{N}_{3 \text { mat, }}$, gives $\mathrm{y}=0.169 \mathrm{x}+1.0 ; \mathrm{R}^{2}=0.40 ; \mathrm{p}<0.05$. The two regression plots are shown in Figure 5.5.2.1.

The WG decided that the February-March 2002 estimate of the abundance of immature 2-group capelin (year class 2001) was a reasonable basis for predicting the abundance of maturing capelin of the 2001 year class on 1 August 2003 and the estimate of the 2000 year class can be predicted from the historic relationship between the total abundance of
each year class at age 3 and the contribution of the same year class at age 4 . To correspond to the timeline used in the past, these data were projected back in time to November 2002 using a monthly $\mathrm{M}=0.035$. The resulting projections of 63.0 and 15.6 billion mature fish belonging to the 2001 and 2000 year classes respectively, are given in Table 5.5.1.1.

During the last ten years, the weight-at-age of adult capelin has been inversely related to adult stock abundance by simple linear regressions having $\mathrm{r}^{2}=0.64$ and 0.68 for age groups 2 and 3 , respectively. These two regression plots are shown in Figure 5.5.2.2. Applying the appropriate regression equations; $y=-0.034 x+19.4 ; r^{2}=0.64 ; p<0.05$ for the younger component, and $y=-0.068 x+29.0 ; r^{2}=0.68 ; p<0.05$ for the older one and using the predicted abundance of age groups 2 and 3 on 1 August 2003 combined, i.e. $78.6 * 10^{9}$ fish, results in estimated mean weights of 16.7 and 23.9 $g$ for age groups 2 and 3 , respectively.

Applying the estimated mean weight, results in a predicted TAC of 835000 t spread evenly from August 2003-March 2004. This corresponds to a preliminary TAC of 555000 t . As in previous years, decisions on the final TAC for the 2003/2004 season should be based on surveys carried out in October/November 2003 and January/February 2004.

### 5.5.3 Management of capelin in the Iceland-East Greenland-Jan Mayen area

The fishable stock consists of two age groups (2-and 3-year-olds spawning at ages 3 and 4). The fishing season usually begins in June/July and ends in March of the following year when the remainder of the fishable stock spawns and dies. The fishable stock, which is also the maturing stock, is renewed annually and its exploitation must of necessity be cautious. Due to the short life span and high spawning mortality, stock abundance can only be assessed by acoustic surveys.

Since 1992, the key elements in the management of capelin in the Iceland-East Greenland-Jan Mayen area have been as follows:

Acoustic survey estimates of juvenile capelin abundance have been used to predict fishable stock abundance in the following year (fishing season). Historical average mean weight-at-age (in later years a relationship between numerical stock abundance and growth), growth rates and natural mortality have been used for calculations and projections of maturing and fishable stock biomass.

Based on the data described above, a TAC is predicted in the spring of the year in which the season begins, allowing for 400000 t to spawn at the end of the season. For precautionary purposes, a preliminary TAC, corresponding to $2 / 3$ of the predicted total TAC for the season, has tbeen allocated to the period July-December. With regard to a precautionary approach, the WG stresses the importance of the continued setting of a preliminary TAC for the first half of the season.

The final decision on a TAC for each fishing season has been based on the results of acoustic abundance surveys in late autumn or in January/February of the following year during the fishing season.

The procedure just described has worked well in the past for 'normal' ranges of stock abundance. However, it is clear that extra care should be taken when dealing with stock abundances below or above the norm, corresponding to TACs lower than 500000 t or greater than 1600000 t .

### 5.6 Precautionary Approach to Fisheries Management

Due to the short life span of capelin and their high spawning mortality, the main management objective is to maintain enough spawners for the propagation of the stock. Since 1979 the targeted remaining spawning stock for capelin in the Iceland-East Greenland Jan Mayen area has been 400000 t . Although there have been large fluctuations in stock abundance during this period, these appear to be environmentally induced and not the result of excessive fishing. Therefore, the criterion of maintaining a remaining spawning stock may be defined as $\mathrm{B}_{\mathrm{lim}}$, i.e. stock abundance below which no fishery should be permitted.

The definition of other precautionary reference points is more problematic. However, due to uncertainties inherent in predicting the abundance of short-lived species and the importance of capelin as a forage fish for predators such as cod, saithe, Greenland halibut, baleen whales and sea birds, extra precaution should be taken when stock biomass projections indicate TACs lower than 500000 t and greater than 1600000 t . In the former case, the fishery should not be opened until the completion of a stock assessment survey in autumn/winter in that season. The latter simply represents a scenario where projected stock abundance is beyond the highest historical abundance on record. In such cases the preliminary TAC should not exceed 1100000 t .

A new model for predicting fishable stock abundance was developed in 2001 (Gudmundsdottir and Vilhjalmsson 2002; see also section 5.7.2). The main advantage of this model is that it appears to predict the abundance of the older age group more precisely than the model presently used. This is especially true for periods of high maturing ratios of age group 2 to spawn at age 3 , which results in low numbers of the same year class spawning at age 4 , The danger of overestimates of fishable stock biomass in the coming is thus alleviated. In the present case, this model predicts total adult capelin biomass (TSB) of 1305000 in summer 2003 and a TAC of 742000 t for all of the 2003/04 season. Corresponding predictions by the current. model are 1425000 t and 835000 t for TSB and TAC, respectively.

### 5.7 Special Comments

### 5.7.1 Fishing restrictions

In most years, the largest capelin can be caught in late June, July and the first half of August. After that, the average size in the catches has usually declined drastically and not increased again until late autumn. There are two main reasons for this. First, the oldest and largest fish migrate ahead of other stock components to feed in the plankton-rich area between Iceland, Greenland and Jan Mayen. Later on, these larger capelin are joined by younger, slower growing adults and even juveniles in parts of the fishing area, their location being variable from year to year. Second, as the food supply diminishes in the southern part of the feeding area in August, the fishable stock becomes more scattered and sometimes mixes with juveniles.

The WG recommends that the 2003 summer/autumn season be opened around 20 June. To prevent catches of juvenile capelin (ages 1 and 2) it is recommended that the authorities responsible for the management of this stock (Greenland, Iceland and Norway) monitor the fishery and be prepared to intervene quickly on short notice using area closures to prevent fishing on mixed concentrations of juveniles and adults.

An overview of stock development during 1978-2002 is given in Table 5.7.1.

### 5.7.2 Research, exploitation and ecological considerations

Capelin is by far the most important forage fish species in the marine ecosystem around Iceland, as well as in the Denmark Strait and the Iceland Sea, where adult capelin feed in summer. Furthermore, these capelin have been the subject of a large fishery which began in a modest way in winter 1964. The yield of the winter fishery gradually increased to about 500000 t in the mid-1970s. A summer fishery began in 1976 and in 1978 the annual catch rose to more than 1150000 t . The stock all but collapsed in the early 1980 s , but recovered after a fishing ban of two years. Since 1983 annual catches have varied between about 260000 t and 1560000 t with an average of 1020000 t .

The large catches, taken in summer and autumn of 1980 and 1981, brought the spawning stock down to very low levels and appear to have caused a severe reduction ( $25-30 \%$ ) of the mean weight of medium size cod (ages 5-8 years). Nevertheless, fairly good year classes recruited from these small spawning stocks of capelin and exploitation-related accidents have not occurred since then. However, mistakes in assessments and the subsequent management of such short-lived species will inevitably be made at some point(s) of time in future and could even be of such magnitude that the stock is severely depleted or, in a worst-case scenario, fished out. For this reason it is imperative that possible weaknesses in the set of data, collected in the past as well as in future, be identified and the problems addressed and resolved when possible. At present the most obvious amendments needed are as follows:

1) Since 1997, the feeding area of the maturing stock has shifted westward and not been as extensive geographically as it was in the late 1970s and during the 1980s. In this period, return migrations, from the northern feeding areas to the outer parts and the slope of the north Icelandic shelf, have been progressively delayed and the capelin stayed farther offshore during December-February than in most years prior to 1997. Furthermore, the capelin have, as a rule, scattered in late July and not formed schools of fishable densities until 10-20 December instead of October-November in the earlier period. In the last four years the 'routine' acoustic/trawl abundance surveys, carried out in November, have missed most of the adult fishable stock, and in November 2003 the autumn juvenile assessment also failed. Although these variations are almost certainly caused by environmental changes, there are practically no available data with which to investigate the causes of the distribution and behaviour changes just described. Because of the dual role of capelin as a highly important prey item in the marine ecosystem and an object of a large directed fishery, the importance of filling this gap in knowledge is obvious and imperative. As described above, the failure of assessing the juvenile stock component and, hence, correctly predicting fishable stock abundance and TACs, can have very serious consequences, both for the capelin and its dependants.
2) The model, presently used to predict the abundance of this capelin stock from historical relationships between the numbers measured as 1 -year-olds by acoustics in autumn, and the back-calculated numerical abundance of
the same year classes which actually matured to spawn at age 3. i.e. about 16 months later, has worked well in the past. However, it has proven difficult to accurately predict the abundance of that part of each year class which does not mature and spawn until at age 4 . In this case a comparison is made between the back-calculated abundance of capelin which spawned at age 4 , on one hand, and that of the same year class which spawned at age 3 plus the numbers of the same year class measured by acoustics as immature 3 -year-olds, on the other. A new prediction model was made in 2001 (Gudmundsdottir and Vilhjalmsson 2002), based on a similar comparison between acoustic estimates and historic data for both age groups, but using log-transformed data from a shorter time-series for the older one. When tested on historical data both models show similar trends. However, the new model gives more conservative predictions for the older year class, has considerably narrower confidence limits and is less cumbersome to comprehend and use. The WG did not evaluate the pros and cons of both models but will do so in 2004.
3) In the past, there have at times been substantial deviations of ageing between Icelandic and Norwegian readers. During periods of intensive fishing and high catch rates, such differences can cause serious bias and confound retrospective assessments of year class size, on which projections of future stock developments are based. This problem must be addressed. An international workshop on the ageing of capelin from all areas where they occur and are studied, is being planned. The the exact location and date have not yet been decided.

| Investigation | No. of samples | Length meas. individuals | Aged individuals |
| :--- | :---: | :---: | :---: |
| Fishery 2002 | 13 | 2700 | 1235 |
| Survey 2002 | 15 | 1500 | 1465 |
| Fishery 2003 | 36 | 3600 | 3587 |
| Survey 2003 | 65 | 6500 | 5415 |

Table 5.1.1 Preliminary TACs for the summer/autumn fishery, recommended TACs for the entire season, landings and remaining spawning stock (000 tonnes) in the 1990/91-2002/03 seasons.

| Season | $90 / 91$ | $91 / 92$ | $92 / 93$ | $93 / 94$ | $94 / 95$ | $95 / 96$ | $96 / 97$ | $97 / 98$ | $98 / 99$ | $99 / 00$ | $00 / 01$ | $01 / 02$ | $02 / 03$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Prelim. TAC | 600 | 0 | 500 | 900 | 950 | 800 | 1100 | 850 | 950 | 866 | 975 | 1050 | 1040 |
| Rec. TAC | 250 | 740 | 900 | 1250 | 850 | 1390 | 1600 | 1265 | 1200 | 1000 | 1090 | 1325 | 1000 |
| Landings | 314 | 677 | 788 | 1179 | 842 | 930 | 1571 | 1245 | 1100 | 934 | 1065 | 1249 | 988 |
| Spawn. stock | 330 | 475 | 460 | 460 | 420 | 830 | 430 | 492 | 500 | 650 | 450 | 475 | 410 |

Table 5.2.1 The international capelin catch 1964-2003 (thousand tonnes).

| Year | Winter season |  |  |  |  | Summer and autumn season |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Iceland | Norway | Faroes | Greenland | $\begin{array}{r} \hline \text { Season } \\ \text { total } \end{array}$ | Iceland | Norway | Faroes | Greenland | EU | $\begin{array}{r} \text { Season } \\ \text { total } \end{array}$ |  |
| 1964 | 8.6 | - | - |  | 8.6 | - | - | - |  | - | - | 8.6 |
| 1965 | 49.7 | - | - |  | 49.7 | - | - | - |  | - | - | 49.7 |
| 1966 | 124.5 | - | - |  | 124.5 | - | - | - |  | - | - | 124.5 |
| 1967 | 97.2 | - | - |  | 97.2 | - | - | - |  | - | - | 97.2 |
| 1968 | 78.1 | - | - |  | 78.1 | - | - | - |  | - | - | 78.1 |
| 1969 | 170.6 | - | - |  | 170.6 | - | - | - |  | - | - | 170.6 |
| 1970 | 190.8 | - | - |  | 190.8 | - | - | - |  | - | - | 190.8 |
| 1971 | 182.9 | - | - |  | 182.9 | - | - | - |  | - | - | 182.9 |
| 1972 | 276.5 | - | - |  | 276.5 |  | - | - |  | - | - | 276.5 |
| 1973 | 440.9 | - | - |  | 440.9 | - | - | - |  | - | - | 440.9 |
| 1974 | 461.9 | - | - |  | 461.9 | - | - | - |  | - | - | 461.9 |
| 1975 | 457.1 | - | - |  | 457.1 | 3.1 | - | - |  | - | 3.1 | 460.2 |
| 1976 | 338.7 | - | - |  | 338.7 | 114.4 | - | - |  | - | 114.4 | 453.1 |
| 1977 | 549.2 | - | 24.3 |  | 573.5 | 259.7 | - | - |  | - | 259.7 | 833.2 |
| 1978 | 468.4 | - | 36.2 |  | 504.6 | 497.5 | 154.1 | 3.4 |  | - | 655.0 | 1,159.6 |
| 1979 | 521.7 | - | 18.2 |  | 539.9 | 442.0 | 124.0 | 22.0 |  | - | 588.0 | 1,127.9 |
| 1980 | 392.1 | - | - |  | 392.1 | 367.4 | 118.7 | 24.2 |  | 17.3 | 527.6 | 919.7 |
| 1981 | 156.0 | - | - |  | 156.0 | 484.6 | 91.4 | 16.2 |  | 20.8 | 613.0 | 769.0 |
| 1982 | 13.2 | - | - |  | 13.2 | - | - | - |  | - | - | 13.2 |
| 1983 | - | - | - |  | - | 133.4 | - | - |  | - | 133.4 | 133.4 |
| 1984 | 439.6 | - | - |  | 439.6 | 425.2 | 104.6 | 10.2 |  | 8.5 | 548.5 | 988.1 |
| 1985 | 348.5 | - | - |  | 348.5 | 644.8 | 193.0 | 65.9 |  | 16.0 | 919.7 | 1,268.2 |
| 1986 | 341.8 | 50.0 | - |  | 391.8 | 552.5 | 149.7 | 65.4 |  | 5.3 | 772.9 | 1,164.7 |
| 1987 | 500.6 | 59.9 | - |  | 560.5 | 311.3 | 82.1 | 65.2 |  | - | 458.6 | 1,019.1 |
| 1988 | 600.6 | 56.6 | - |  | 657.2 | 311.4 | 11.5 | 48.5 |  | - | 371.4 | 1,028.6 |
| 1989 | 609.1 | 56.0 | - |  | 665.1 | 53.9 | 52.7 | 14.4 |  | - | 121.0 | 786,1 |
| 1990 | 612.0 | 62.5 | 12.3 |  | 686,8 | 83.7 | 21.9 | 5.6 |  | - | 111.2 | 798.0 |
| 1991 | 202.4 | - | - |  | 202.4 | 56.0 | - | - |  | - | 56.0 | 258.4 |
| 1992 | 573.5 | 47.6 | - |  | 621.1 | 213.4 | 65.3 | 18.9 | 0.5 |  | 298.1 | 919.2 |
| 1993 | 489.1 | - | - | 0.5 | 489.6 | 450.0 | 127.5 | 23.9 | 10.2 |  | 611.6 | 1,101.2 |
| 1994 | 550.3 | 15.0 | - | 1.8 | 567.1 | 210.7 | 99.0 | 12.3 | 2.1 |  | 324.1 | 891.2 |
| 1995 | 539.4 | - | - | 0.4 | 539.8 | 175.5 | 28.0 | - | 2.2 |  | 205.7 | 745.5 |
| 1996 | 707.9 | - | 10.0 | 5.7 | 723.6 | 474.3 | 206.0 | 17.6 | 15.0 | 60.9 | 773.8 | 1,497.4 |
| 1997 | 774.9 | - | 16.1 | 6.1 | 797.1 | 536.0 | 153.6 | 20.5 | 6.5 | 47.1 | 763.6 | 1,561.5 |
| 1998 | 457.0 | - | 14.7 | 9.6 | 481.3 | 290.8 | 72.9 | 26.9 | 8.0 | 41.9 | 440.5 | 921.8 |
| 1999 | 607.8 | 14.8 | 13.8 | 22.5 | 658.9 | 83.0 | 11.4 | 6.0 | 2.0 |  | 102.4 | 761.3 |
| 2000 | 761.4 | 14.9 | 32.0 | 22.0 | 830.3 | 126.5 | 80.1 | 30.0 | 7.5 | 21.0 | 265.1 | 1,095.4 |
| 2001 | 767.2 | - | 10.0 | 29.0 | 806.2 | 150.0 | 106.0 | 12.0 | 9.0 | 17.0 | 294.0 | 1,061.2 |
| 2002 | 901.0 | - | 28.0 | 26.0 | 955.0 | 180.0 | 118.7 | - | 13.0 | 28.0 | 339.7 | 1,294.7 |
| 2003 | 585.0 | - | 40.0 | 23.0 | 648.0 |  |  |  |  |  |  |  |

Table 5.2.2 The total international catch of capelin in the Iceland-East Greenland-Jan Mayen area by age group in numbers (billions) and the total catch by numbers and weight (thousand tonnes) in the autumn season (August-December) 1981-2002.

|  |  |  | Year |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 1 | 0.6 | - | 0.6 | 0.5 | 0.8 | + | + | 0.3 | 1.7 | 0.8 | 0.3 |
| 2 | 27.9 | - | 7.2 | 9.8 | 25.6 | 10.0 | 27.7 | 13.6 | 6.0 | 5.9 | 2.7 |
| 3 | 2.0 | - | 0.8 | 7.8 | 15.4 | 23.3 | 6.7 | 5.4 | 1.5 | 1.0 | 0.4 |
| 4 | + | - | - | 0.1 | 0.2 | 0.5 | + | + | + | + | + |
| Total number | 30.5 | - | 8.6 | 18.2 | 42.0 | 33.8 | 34.4 | 19.3 | 9.2 | 7.7 | 3.4 |
| Total weight | 613.0 | - | 133.4 | 548.5 | 919.7 | 772.9 | 458.6 | 371.4 | 121.0 | 111.2 | 56.0 |


|  |  |  |  | Year |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| 1 | 1.7 | 0.2 | 0.6 | 1.5 | 0.2 | 1.8 | 0.9 | 0.3 | 0.2 | + | + |
| 2 | 14.0 | 24.9 | 15.0 | 9.7 | 25.2 | 33.4 | 25.1 | 4.7 | 12.9 | 17.6 | 18.3 |
| 3 | 2.1 | 5.4 | 2.8 | 1.1 | 12.7 | 10.2 | 2.9 | 0.7 | 3.3 | 1.2 | 2.5 |
| 4 | + | 0.2 | + | + | 0.2 | 0.4 | + | + | 0.1 | + | + |
| Total number | 17.8 | 30.7 | 18.4 | 12.3 | 38.4 | 45.8 | 28.9 | 5.7 | 16.5 | 18.8 | 20.8 |
| Total weight | 298.1 | 611.6 | 324.1 | 205.7 | 773.7 | 763.6 | 440.5 | 102.4 | 265.1 | 294.0 | 339.7 |

Table 5.2.3 The total international catch of capelin in the Iceland-East Greenland-Jan Mayen area by age group in numbers (billions) and the total catch by numbers and weight (thousand tonnes) in the winter season (January-March) 1982-2003.

|  |  |  | Year |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| 2 | - | - | 2.1 | 0.4 | 0.1 | + | + | 0.1 | 1.4 | 0.5 | 2.7 |
| 3 | 0.8 | - | 18.1 | 9.1 | 9.8 | 6.9 | 23.4 | 22.9 | 24.8 | 7.4 | 29.4 |
| 4 | 0.1 | - | 3.4 | 5.4 | 6.9 | 15.5 | 7.2 | 7.8 | 9.6 | 1.5 | 2.8 |
| 5 | - | - | - | - | 0.2 | - | 0.3 | + | 0.1 | + | + |
| Total number | 0.9 | - | 23.6 | 14.5 | 17.0 | 22.4 | 30.9 | 30.8 | 35.9 | 9.4 | 34.9 |
| Total weight | 13.2 | - | 439.6 | 348.5 | 391.8 | 560.5 | 657.2 | 665.1 | 686.8 | 202.4 | 621.1 |


|  |  | Year |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| 2 | 0.2 | 0.6 | 1.3 | 0.6 | 0.9 | 0.3 | 0.5 | 0.3 | 0.4 | 0.1 | 0.1 |
| 3 | 20.1 | 22.7 | 17.6 | 27.4 | 29.1 | 20.4 | 31.2 | 36.3 | 27.9 | 33.1 | 32.2 |
| 4 | 2.5 | 3.9 | 5.9 | 7.7 | 11.0 | 5.4 | 7.5 | 5.4 | 6.7 | 4.2 | 1.9 |
| 5 | + | + | + | + | + | + | + | + | + | + | + |
| Total number | 22.8 | 27.2 | 24.8 | 35.7 | 41.0 | 26.1 | 39.2 | 42.0 | 35.0 | 37.4 | 34.4 |
| Total weight | 489.6 | 567.1 | 539.8 | 723.6 | 797.6 | 481.3 | 658.9 | 830.3 | 787.2 | 955.0 | 648.0 |

Table 5.2.4 The total international catch in numbers (millions) of capelin in the Iceland-East GreenlandJan Mayen area in the summer/autumn season of 2002 by age and length, and the catch in weight (thousand tonnes) by age group.

| Total length (cm) | Age 1 | Age 2 | Age 3 | Age 4 | Total | Percentage |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 10.5 | 13 | 134 | 0 | 0 | 147 | 0.7 |
| 11 | 21 | 428 | 0 | 0 | 449 | 2.2 |
| 11.5 | 12 | 1783 | 0 | 0 | 1795 | 8.6 |
| 12 | 0 | 3213 | 39 | 0 | 3252 | 15.7 |
| 12.5 | 0 | 2413 | 100 | 0 | 2513 | 12.1 |
| 13 | 0 | 2227 | 250 | 0 | 2477 | 11.9 |
| 13.5 | 0 | 2463 | 211 | 0 | 2675 | 12.9 |
| 14 | 0 | 1843 | 736 | 0 | 2580 | 12.4 |
| 14.5 | 0 | 2046 | 588 | 0 | 2634 | 12.7 |
| 15 | 0 | 646 | 579 | 0 | 1225 | 5.9 |
| 15.5 | 0 | 294 | 300 | 0 | 594 | 2.9 |
| 16 | 0 | 116 | 150 | 0 | 266 | 1.3 |
| 16.5 | 0 | 0 | 100 | 2 | 102 | 0.5 |
| 17 | 0 | 0 | 24 | 5 | 29 | 0.1 |
| 17.5 | 0 | 0 | 15 | 6 | 21 | 0.1 |
| Total number | 46 | 17607 | 3093 | 13 | 20759 |  |
| Percentage | 0.2 | 84.8 | 14.9 | 0.1 | 100 | 100.0 |
| Total weight | 0.3 | 277.1 | 59.7 | 0.3 | 339.7 |  |

Table 5.2.5 The total international catch in numbers (millions) of capelin in the Iceland-East GreenlandJan Mayen area in the winter season of 2003 by age and length, and the catch in weight (thousand tonnes) by age group.

| Total length (cm) | Age 2 | Age 3 | Age 4 | Age 5 | Total | Percentage |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |
| 10.5 | 5 | 4 | 0 | - | 9 | + |  |
| 11 | 3 | 6 | 0 | - | 9 | + |  |
| 11.5 | 1 | 8 | 0 | - | 9 | + |  |
| 12 | - | 38 | 0 | - | 47 | 0.1 |  |
| 12.5 | - | 132 | 0 | - | 132 | 0.4 |  |
| 13 | - | 820 | 9 | - | 829 | 2.5 |  |
|  | 13.5 | - | 1631 | 9 | - | 1640 | 4.9 |
| 14 | - | 2875 | 19 | - | 2894 | 8.6 |  |
|  | 14.5 | - | 4883 | 85 | - | 4967 | 14.7 |
|  | 15 | - | 5514 | 94 | - | 5608 | 16.6 |
|  | 15.5 | - | 5787 | 302 | - | 6089 | 18.0 |
|  | 16 | - | 4996 | 320 | - | 5316 | 15.7 |
|  | 16.5 | - | 3356 | 415 | - | 3770 | 11.2 |
|  | 17 | - | 1348 | 330 | - | 1678 | 5.0 |
|  | 17.5 | - | 339 | 254 | - | 594 | 1.8 |
|  | 18 | - | 94 | 75 | - | 170 | 0.5 |
|  | 18.5 | - | 19 | 0 | - | 19 | 0.1 |
|  |  |  |  |  |  |  |  |
| Total number |  | 9 | 31859 | 1913 | - | 33782 |  |
| Percentage |  | + | 94.3 | 5.7 | - | 100.0 | 100.0 |
| Total weight |  | 0.1 | 601.6 | 46.3 | - | 648.0 |  |

Table 5.3.1.1 Abundance indices of 0-group capelin 1970-2002 and their division by areas.

|  | Year |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 |
| NW-Irminger Sea | 1 | + | + | 14 | 26 | 3 | 2 | 2 | + | 4 | 3 | 10 | + |
| W-Iceland | 8 | 7 | 30 | 39 | 44 | 37 | 5 | 19 | 2 | 19 | 18 | 13 | 8 |
| N-Iceland | 2 | 12 | 52 | 46 | 57 | 46 | 10 | 19 | 29 | 25 | 19 | 6 | 5 |
| East Iceland | - | + | 7 | 17 | 7 | 3 | 15 | 3 | + | 1 | + | - | + |
| Total | 11 | 19 | 89 | 116 | 134 | 89 | 32 | 43 | 31 | 49 | 40 | 29 | 13 |
|  |  |  |  |  |  |  | Year |  |  |  |  |  |  |
| Area | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| NW-Irminger Sea | + | + | 1 | + | 1 | 3 | 1 | + | 8 | 3 | 2 | 3 | + |
| W-Iceland | 3 | 2 | 8 | 16 | 6 | 22 | 13 | 7 | 2 | 11 | 21 | 12 | 6 |
| N -Iceland | 18 | 17 | 19 | 17 | 6 | 26 | 24 | 12 | 43 | 20 | 13 | 69 | 10 |
| East Iceland | 1 | 9 | 3 | 4 | 1 | 1 | 2 | 2 | 1 | + | 15 | 10 | 8 |
| Total | 22 | 28 | 31 | 37 | 14 | 52 | 40 | 21 | 54 | 34 | 51 | 94 | 24 |
|  |  |  |  |  |  |  | Year |  |  |  |  |  |  |
|  | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |  |  |  |  |  |  |
| NW-Irminger Sea | 2 | 5 | + | NA | NA | NA | NA |  |  |  |  |  |  |
| W-Iceland | 17 | 14 | 7 | 25 | 1 | 25 | 17 |  |  |  |  |  |  |
| N -Iceland | 57 | 30 | 34 | 51 | 7 | 53 | 8 |  |  |  |  |  |  |
| East Iceland | 6 | 12 | 5 | 7 | 4 | 4 | 1 |  |  |  |  |  |  |
| Total | 82 | 61 | 46 | 83 | 12 | 82 | 26 |  |  |  |  |  |  |

Table 5.3.1.2 Estimated numbers, mean length and weight of age 1 capelin in the August surveys for 1983-2002.

|  | Year |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| Number ( $10{ }^{9}$ ) | 155 | 286 | 31 | 71 | 101 | 147 | 111 | 36 | 50 | 87 | 33 | 85 | 189 | 138 |
| Mean length (cm) | 10.4 | 9.7 | 10.2 | 9.5 | 9.1 | 8.8 | 10.1 | 10.4 | 10.7 | 9.7 | 9.4 | 9.0 | 9.8 | 9.3 |
| Mean weight (g) | 4.2 | 3.6 | 3.8 | 3.3 | 3.0 | 2.6 | 3.4 | 4.0 | 5.1 | 3.4 | 3.0 | 2.8 | 3.4 | 2.9 |
|  | Year |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |  |  |  |  |  |  |  |  |
| Number ( $10^{9}$ ) | 143 | 87 | 55 | 94 | 99 | No survey |  |  |  |  |  |  |  |  |
| Mean length (cm) | 9.3 | 9.0 | 9.5 | 9.5 | 10.0 |  |  |  |  |  |  |  |  |  |
| Mean weight (g) | 2.8 | 2.9 | 3.2 | 3.1 | 3.7 |  |  |  |  |  |  |  |  |  |

Table 5.3.2.1 Acoustic abundance estimate of maturing capelin, 7-22 January 2003.

| Length (cm) | NUMBERS (109) |  |  |  | Avg <br> wt <br> (g) | BIOMASS ( $10^{-9}$ ) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | (Age) Year class |  |  | Total |  | (Age) Year class |  |  | Total |
|  | (2) 2001 | (3) 2000 | (4) 1999 |  |  | (2) 2001 | (3) 2000 | (4) 1999 |  |
| 12.5 | 0.0 | 0.0 | 0.0 | 0.0 | 6.9 | 0.0 | 0.0 | 0.0 | 0.0 |
| 13 | 0.0 | 100.0 | 0.0 | 100.0 | 8.2 | 0.0 | 0.8 | 0.0 | 0.8 |
| 13.5 | 0.0 | 900.0 | 0.0 | 900.0 | 9.4 | 0.0 | 8.5 | 0.0 | 8.5 |
| 14 | 0.0 | 2700.0 | 0.0 | 2700.0 | 11.0 | 0.0 | 29.7 | 0.0 | 29.7 |
| 14.5 | 0.0 | 5000.0 | 100.0 | 5000.0 | 12.7 | 0.0 | 63.5 | 1.3 | 63.5 |
| 15 | 0.0 | 6400.0 | 200.0 | 6600.0 | 14.4 | 0.0 | 92.2 | 2.9 | 95.0 |
| 15.5 | 0.0 | 8400.0 | 300.0 | 8700.0 | 16.3 | 0.0 | 136.9 | 4.9 | 141.8 |
| 16 | 0.0 | 7200.0 | 300.0 | 7400.0 | 18.7 | 0.0 | 134.6 | 5.6 | 138.4 |
| 16.5 | 0.0 | 5200.0 | 300.0 | 5600.0 | 21.1 | 0.0 | 109.7 | 6.3 | 118.2 |
| 17 | 0.0 | 3900.0 | 600.0 | 4500.0 | 23.6 | 0.0 | 92.0 | 14.2 | 106.2 |
| 17.5 | 0.0 | 2400.0 | 400.0 | 2800.0 | 26.6 | 0.0 | 63.8 | 10.6 | 74.5 |
| 18 | 0.0 | 700.0 | 100.0 | 900.0 | 29.4 | 0.0 | 20.6 | 2.9 | 26.5 |
| 18.5 | 0.0 | 200.0 | 100.0 | 400.0 | 31.7 | 0.0 | 6.3 | 3.2 | 12.7 |
| 19 | 0.0 | 0.0 | 0.0 | 0.0 | 36.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Total | 100.0 | 43100.0 | 2400.0 | 45600.0 |  | 0.0 | 758.7 | 51.9 | 815.7 |
| Average length |  |  |  |  |  | 13.4 | 15.7 | 16.7 | 15.7 |
| Average weight |  |  |  |  |  | 9.3 | 17.7 | 22.3 | 17.9 |

Table 5.3.2.2 Acoustic abundance estimate of maturing capelin, 17 February-13 March 2003.

| Length <br> (cm) | NUMBERS ( $10^{-9}$ ) |  |  |  | Avg wt (g) | BIOMASS ( $10^{-9}$ ) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | (Age) Year class |  |  | Total |  | (Age) Year class |  |  | Total |
|  | (2) 2001 | (3) 2000 | (4) 1999 |  |  | (2) 2001 | (3) 1999 | (4) 1998 |  |
| 13 | 10.5 | 137.1 | 0.0 | 147.6 | 9.6 | 1.3 | 0.0 | 1.4 | 1.4 |
| 13.5 | 10.5 | 482.9 | 0.0 | 493.4 | 10.4 | 5.0 | 0.0 | 5.1 | 5.1 |
| 14 | 0.0 | 1730.9 | 0.0 | 1730.9 | 11.7 | 20.3 | 0.0 | 20.3 | 20.3 |
| 14.5 | 0.0 | 2927.4 | 18.7 | 2946.1 | 13.3 | 38.9 | 0.2 | 39.2 | 39.2 |
| 15 | 0.0 | 4213.3 | 19.9 | 4233.2 | 14.9 | 62.9 | 0.3 | 63.1 | 63.1 |
| 15.5 | 0.0 | 4619.6 | 42.4 | 4662.1 | 16.9 | 78.0 | 0.7 | 78.7 | 78.7 |
| 16 | 0.0 | 4842.4 | 22.5 | 4864.9 | 19.1 | 92.5 | 0.4 | 92.9 | 92.9 |
| 16.5 | 0.0 | 3943.0 | 502.4 | 4445.3 | 21.8 | 86.1 | 11.0 | 97.1 | 97.1 |
| 17 | 0.0 | 2745.5 | 75.9 | 2821.4 | 26.3 | 72.3 | 2.0 | 74.3 | 74.3 |
| 17.5 | 0.0 | 1385.1 | 80.2 | 1465.3 | 27.2 | 37.7 | 2.2 | 39.9 | 39.9 |
| 18 | 0.0 | 374.7 | 112.9 | 487.6 | 29.1 | 10.9 | 3.3 | 14.2 | 14.2 |
| 18.5 | 0.0 | 29.7 | 0.0 | 29.7 | 30.3 | 0.9 | 0.0 | 0.9 | 0.9 |
| 19 | 0.0 | 20.7 | 0.0 | 20.7 | 33.4 | 0.7 | 0.0 | 0.7 | 0.7 |
| Total | 21.0 | 27452.4 | 874.9 | 28348.3 | 18.6 | 507.5 | 20.1 | 527.9 | 527.9 |
| Average length |  |  |  |  |  | 13.3 | 15.7 | 16.7 | 15.7 |
| Average weight |  |  |  |  |  | 10.0 | 18.5 | 23.0 | 18.6 |

Table 5.3.2.3 Acoustic estimate of immature capelin, 13 February-17 March and 4-11 April 2003

| Length (cm) | NUMBERS ( $10^{-9}$ ) |  |  | Avg <br> Wt <br> (g) | BIOMASS $\left(10^{-6} \mathrm{t}\right)$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | (Age) Year class |  | Total |  | (Age) Year class |  | Total |
|  | (2) 1999 | (3) 1998 |  |  | (2) 2001 | (3) 2000 |  |
| 7.5 | 0.916 | 0.000 | 0.916 | 1.5 | 1.3 | 0.0 | 1.3 |
| 8 | 2.565 | 0.000 | 2.565 | 1.7 | 4.3 | 0.0 | 4.3 |
| 8.5 | 3.664 | 0.000 | 3.664 | 2.1 | 7.5 | 0.0 | 7.5 |
| 9 | 1.740 | 0.000 | 1.740 | 2.4 | 4.2 | 0.0 | 4.2 |
| 9.5 | 3.666 | 0.000 | 3.666 | 2.8 | 10.4 | 0.0 | 10.4 |
| 10 | 5.595 | 0.000 | 5.595 | 3.3 | 18.6 | 0.0 | 18.6 |
| 10.5 | 6.509 | 0.000 | 6.509 | 3.8 | 24.4 | 0.0 | 24.4 |
| 11 | 7.636 | 0.000 | 7.636 | 4.4 | 33.3 | 0.0 | 33.3 |
| 11.5 | 7.222 | 0.007 | 7.229 | 5.0 | 36.3 | 0.0 | 36.4 |
| 12 | 5.626 | 0.103 | 5.729 | 5.9 | 33.0 | 0.6 | 33.6 |
| 12.5 | 5.850 | 0.045 | 5.895 | 6.9 | 40.6 | 0.3 | 40.9 |
| 13 | 4.955 | 0.090 | 5.045 | 7.8 | 38.7 | 0.7 | 39.4 |
| 13.5 | 2.499 | 0.552 | 3.051 | 9.0 | 22.5 | 5.0 | 27.4 |
| 14 | 2.294 | 0.963 | 3.257 | 10.4 | 23.8 | 10.0 | 33.8 |
| 14.5 | 0.824 | 0.940 | 1.765 | 12.1 | 10.0 | 11.4 | 21.3 |
| 15 | 0.824 | 0.866 | 1.690 | 12.7 | 10.5 | 11.0 | 21.4 |
| 15.5 | 0.092 | 0.150 | 0.242 | 14.7 | 1.3 | 2.2 | 3.6 |
| Total | 62.476 | 3.716 | 66.192 | 5.5 | 341.8 | 20.3 | 362.1 |
| Average length (cm) |  |  |  |  | 11.2 | 14.2 | 11.3 |
| Average weight (g) |  |  |  |  | 5.1 | 11.1 | 5.5 |

Table 5.4.1 The estimated number (billions) of capelin on 1 August 1978-2003 by age and maturity groups. The total number (billions) and weight (thousand tonnes) of the immature and maturing (fishable) stock components are also given.

|  | Year |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age/maturity | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| 1 juvenile | 163.8 | 60.3 | 66.1 | 48.9 | 146.4 | 124.2 | 250.5 | 98.9 | 156.2 | 144.0 |
| 2 immature | 15.3 | 16.4 | 4.2 | 3.7 | 15.0 | 42.5 | 40.9 | 100.0 | 29.4 | 37.2 |
| 2 mature | 81.9 | 91.3 | 35.4 | 39.7 | 17.1 | 53.7 | 40.7 | 64.6 | 35.6 | 65.4 |
| 3 mature | 29.1 | 10.1 | 10.8 | 2.8 | 2.3 | 9.8 | 27.9 | 27.0 | 65.8 | 20.1 |
| 4 mature | 0.4 | 0.3 | + | + | + | 0.1 | 0.4 | 0.4 | 0.7 | 0.1 |
| Number immat. | 179.2 | 76.7 | 70.3 | 52.6 | 161.4 | 166.7 | 291.4 | 198.9 | 185.6 | 181.2 |
| Number mature | 11.4 | 101.7 | 46.2 | 42.5 | 19.4 | 63.6 | 69.0 | 92.0 | 102.1 | 85.6 |
| Weight immat | 751 | 366 | 283 | 209 | 683 | 985 | 1067 | 1168 | 876 | 950 |
| Weight mature | 2081 | 1769 | 847 | 829 | 355 | 1085 | 1340 | 1643 | 2260 | 1689 |


|  |  |  |  | Year |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age/maturity | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| 1 juvenile | 80.8 | 63.9 | 117.5 | 132.9 | 162.9 | 144.3 | 224.1 | 197.3 | 191.2 | 165.4 |
| 2 immature | 24.0 | 10.3 | 10.1 | 9.7 | 16.6 | 20.1 | 35.2 | 45.1 | 28.7 | 35.2 |
| 2 mature | 70.3 | 42.8 | 31.9 | 67.7 | 70.7 | 86.9 | 59.8 | 102.2 | 100.7 | 90.3 |
| 3 mature | 24.5 | 15.8 | 6.8 | 6.7 | 6.4 | 10.9 | 13.2 | 23.0 | 29.6 | 19.0 |
| 4 mature | 0.4 | + | + | + | + | 0.2 | - | + | + | + |
| Number immat. | 104.8 | 74.2 | 127.6 | 142.6 | 179.5 | 164.7 | 259.2 | 242.4 | 219.9 | 200.6 |
| Number mature | 95.2 | 58.6 | 38.7 | 74.4 | 77.1 | 98.0 | 73.0 | 125.1 | 130.3 | 109.3 |
| Weight immat | 438 | 309 | 542 | 702 | 747 | 702 | 1019 | 1188 | 985 | 758 |
| Weight mature | 1663 | 1173 | 751 | 1273 | 1311 | 1585 | 1268 | 1819 | 1900 | 1590 |


|  |  |  |  | Year |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | :--- |
| Age/maturity | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| 1 juvenile | 167.9 | 138.0 | 140.9 | $110.2^{*}$ | $95.9^{*}$ |  |
| 2 immature | 19.2 | 24.4 | 25.0 | 9.0 | $23.7^{*}$ |  |
| 2 mature | 89.5 | 85.9 | 65.7 | 86.7 | 68.0 | $63.0^{* *}$ |
| 3 mature | 23.2 | 12.6 | 16.0 | 16.9 | 5.9 | $15.6^{* *}$ |
| 4 mature | + | + |  |  |  |  |
| Number immat. | 187.1 | 162.4 | 165.9 | $150.1^{*}$ | $119.6^{*}$ |  |
| Number mature | 112.7 | 98.5 | 81.7 | 103.6 | 73.9 | $78.6^{* *}$ |
| Weight immat | 621 | 612 | 595 | $622^{* 1)}$ | $490^{*}$ |  |
| Weight mature | 1576 | 1703 | 1519 | 1817 | 1280 | $1423^{* *}$ |

* Preliminary
${ }^{* *}$ Predicted

Table 5.4.2 The estimated number (billions) of capelin on 1 January 1979-2003 by age and maturity groups. The total number (billions) and weight (thousand tonnes) of the immature and maturing (fishable) stock components and the remaining spawning stock by number and weight are also given.

|  |  | Year |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age/maturity | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| 2 juvenile | 137.6 | 50.6 | 55.3 | 41.2 | 123.7 | 105.0 | 211.6 | 83.2 | 131.9 | 120.5 |
| 3 immature | 12.8 | 13.8 | 3.5 | 3.0 | 12.6 | 35.7 | 34.3 | 83.9 | 25.6 | 31.2 |
| 3 mature | 51.8 | 53.4 | 16.3 | 8.0 | 14.3 | 39.8 | 25.2 | 34.5 | 22.1 | 34.1 |
| 4 mature | 14.8 | 3.6 | 4.9 | 0.5 | 2.0 | 7.6 | 15.6 | 10.5 | 37.0 | 11.7 |
| 5 mature | 0.3 | 0.2 | + | + | + | 0.1 | 0.3 | 0.2 | 0.2 | + |
| Number immat. | 150.4 | 64.4 | 58.8 | 44.2 | 136.3 | 140.7 | 245.9 | 167.1 | 157.5 | 151.3 |
| Number mature | 66.9 | 57.2 | 21.2 | 8.5 | 16.3 | 47.5 | 41.1 | 45.2 | 59.1 | 45.8 |
| Weight immat. | 1028 | 502 | 527 | 292 | 685 | 984 | 1467 | 1414 | 1003 | 1083 |
| Weight mature | 1358 | 980 | 471 | 171 | 315 | 966 | 913 | 1059 | 1355 | 993 |
| Number sp.st. | 29.0 | 17.5 | 7.7 | 6.8 | 13.5 | 21.6 | 20.7 | 19.6 | 18.3 | 18.5 |
| Weight sp. st | 600 | 300 | 170 | 140 | 260 | 440 | 460 | 460 | 420 | 400 |


|  |  |  | Year |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age/maturity | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| 2 juvenile | 67.8 | 53.9 | 98.9 | 111.6 | 124.6 | 121.3 | 188.1 | 165.2 | 160.0 | 138.8 |
| 3 immature | 20.1 | 8.6 | 8.6 | 8.1 | 13.9 | 16.9 | 29.5 | 37.9 | 24.1 | 29.5 |
| 3 mature | 48.8 | 31.2 | 22.3 | 54.8 | 46.5 | 50.5 | 35.1 | 75.5 | 72.4 | 50.1 |
| 4 mature | 16.0 | 12.1 | 4.5 | 5.3 | 3.5 | 4.6 | 8.7 | 20.1 | 24.8 | 7.9 |
| 5 mature | 0.3 | + | + | + | + | + | + | + | + | + |
| Number immat. | 87.9 | 62.5 | 107.5 | 119.7 | 138.5 | 138.2 | 217.6 | 203.1 | 184.1 | 168.3 |
| Number mature | 64.8 | 43.3 | 26.8 | 60.1 | 50.0 | 55.1 | 43.8 | 95.6 | 97.2 | 58.0 |
| Weight immat. | 434 | 291 | 501 | 487 | 622 | 573 | 696 | 800 | 672 | 621 |
| Weight mature | 1298 | 904 | 544 | 1106 | 1017 | 1063 | 914 | 1820 | 1881 | 1106 |
| Number sp.st. | 22.0 | 5.5 | 16.3 | 25.8 | 23.6 | 24.8 | 19.2 | 42.8 | 21.8 | 27.6 |
| Weight sp. st. | 440 | 115 | 330 | 475 | 499 | 460 | 420 | 830 | 430 | 492 |


|  |  | Year |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Age/maturity | 1999 | 2000 | 2001 | 2002 | 2003 |
| 2 juvenile | 140.9 | 115.8 | 96.9 | $119.2^{*}$ | $80.5^{*}$ |
| 3 immature | 16.1 | 20.5 | 17.2 | $7.6^{*}$ | $19.9^{*}$ |
| 3 mature | 53.2 | 68.2 | 46.3 | 59.3 | 58.4 |
| 4 mature | 16.0 | 10.0 | 10.5 | 10.5 | 2.9 |
| 5 mature | + | + | + | + |  |
| Number immat. | 157.0 | 136.3 | 114.1 | $126.6^{*}$ | $100.4^{*}$ |
| Number mature | 69.3 | 78.2 | 56.8 | $69.8^{*}$ | 61.3 |
| Weight immat. | 585 | 535 | 478 | $510^{*}$ | $575^{*}$ |
| Weight mature | 1171 | 1485 | 1197 | 1445 | 1214 |
| Number sp.st. | 29.5 | 34.2 | 21.3 | 22.9 | 20.7 |
| Weight sp. st. | 500 | 650 | 450 | 475 | 410 |
|  |  |  |  |  |  |
|  |  |  |  |  |  |

Table 5.5.1.1 The data used in the comparisons between abundance of age groups (numbers) when predicting fishable stock abundance for the calculation of preliminary TACs.

|  | Age 1 <br> Acoustics | Age 2 <br> Back-calc. <br> Mature | Age 2 <br> Acoustics <br> Immature | Age 2 <br> Back-calc. <br> Total | Age 3 <br> Back-calc. <br> Mature |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | $\mathrm{N}_{1}$ | $\mathrm{~N}_{2 \text { mat }}$ | $\mathrm{N}_{2 \text { imm }}$ | $\mathrm{N}_{2 \text { tot }}$ | $\mathrm{N}_{3 \text { tot }}$ |
| class | 23.7 | 17.1 | 1.7 | 32.1 | 9.8 |
| 1980 | 68.0 | 53.7 | 8.2 | 96.2 | 27.9 |
| 1981 | 44.1 | 40.7 | 4.6 | 81.6 | 27.0 |
| 1982 | 73.8 | 64.6 | 12.6 | 164.6 | 65.8 |
| 1983 | 33.8 | 35.6 | 1.4 | 65.0 | 20.1 |
| 1984 | 58.0 | 65.4 | 5.4 | 102.6 | 24.5 |
| 1985 | 70.2 | 70.3 | 6.7 | 94.6 | 15.8 |
| 1986 | 43.9 | 42.8 | 1.8 | 53.1 | 6.8 |
| 1987 | 29.2 | 31.9 | 1.3 | 42.0 | 6.7 |
| 1988 | 39.2 | 67.7 | 5.2 | 77.2 | 6.4 |
| 1989 | 60.0 | 70.7 | 2.3 | 87.3 | 10.9 |
| 1990 | 104.6 | 86.9 | 10.8 | 107.0 | 13.2 |
| 1991 | 100.4 | 59.8 | 6.9 | 95.0 | 24.0 |
| 1992 | 119.0 | 102.2 | 46.3 | 147.2 | 29.6 |
| 1993 | 165.0 | 100.7 | 16.4 | 129.4 | 19.0 |
| 1994 | 111.9 | 90.3 | 30.8 | 125.5 | 23.2 |
| 1995 | 128.5 | 89.5 | 6.3 | 108.7 | 12.6 |
| 1996 | 121.0 | 85.9 | 5.0 | 110.3 | 16.0 |
| 1997 | 89.8 | 65.7 | 11.0 | 84.1 | 16.9 |
| 1998 | 103.0 | 86.7 | 2.4 | 95.8 | 8.6 |
| 1999 | 100.3 | 85.4 | 3.7 | $* 89.8$ |  |
| 2000 | 74.4 |  |  |  |  |
| 2001 |  |  |  |  |  |

* Invalid due to ice conditions.
${ }^{* *}$ Preliminary

Table 5.5.1.2 Mean weight $(\mathrm{g})$ in autumn of maturing capelin.

|  |  |  | Years |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| Age 2 | 19.2 | 16.5 | 16.1 | 15.8 | 15.5 | 18.1 | 17.9 | 15.5 |
| Age 3 | 24.0 | 24.1 | 22.5 | 25.7 | 23.8 | 24.1 | 25.8 | 23.4 |
|  |  |  |  |  | Years |  |  |  |
|  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| Age 2 | 18.0 | 18.1 | 16.3 | 16.5 | 16.2 | 16.0 | 15.3 | 15.8 |
| Age 3 | 25.5 | 25.5 | 25.4 | 22.6 | 23.3 | 23.6 | 20.5 | 20.6 |
|  |  |  |  |  |  |  |  |  |
|  | 1997 | 1998 | 1999 | Years |  |  |  |  |
| Age 2 | 14.3 | 14.1 | 16.8 | 17.1 | 16.3 | No data |  |  |
| Age 3 | 20.3 | 18.1 | 20.6 | 24.7 | 23.9 | No data |  |  |

Table 5.5.1.3 Predictions of fishable stock abundance and TACs for the 1984/85-2002/03 seasons.
The last row gives contemporary advice on TACs for comparison.
Age 2 and age $3=$ Numbers in billions in age groups at the beginning of season.
Fishable stock = calculated weight of maturing capelin in thousand tonnes (ref. August).
TAC calc $=$ predicted in thousand tonnes.

| Season | $85 / 86$ | $86 / 87$ | $87 / 88$ | $88 / 89$ | $89 / 90$ | $90 / 91$ | $91 / 92$ | $92 / 93$ | $93 / 94$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year classes | $83-82$ | $84-83$ | $85-84$ | $86-85$ | $87-86$ | $88-87$ | $89-88$ | $90-89$ | $91-90$ |
| Age 2 | 67.8 | 34.9 | 55.5 | 64.8 | 43.2 | 31.1 | 39.4 | 56.4 | 93.1 |
| Age 3 | 20.2 | 55.0 | 13.7 | 29.0 | 25.5 | 8.2 | 3.7 | 18.3 | 22.6 |
| Fishable stock | 1637 | 1926 | 1268 | 1800 | 1350 | 724 | 755 | 1398 | 2123 |
| Calculated TAC | 963 | 1215 | 642 | 1105 | 713 | 170 | 197 | 755 | 1385 |
| Advised TAC | 1311 | 1333 | 1115 | 1036 | 550 | 265 | 740 | $* 900$ | 1250 |
|  |  |  |  |  |  |  |  |  |  |
| Season | $94 / 95$ | $95 / 96$ | $96 / 97$ | $97 / 98$ | $98 / 99$ | $99 / 00$ | $00 / 01$ | $01 / 02$ | $02 / 03$ |
| Year classes | $92-91$ | $93-92$ | $94-93$ | $95-94$ | $96-95$ | $97 / 96$ | $98 / 97$ | $99 / 98$ | $00 / 99$ |
| Age 2 | 89.6 | 92.5 | 90.0 | 83.8 | 94.4 | 89.2 | 70.9 | 77.1 | 77.2 |
| Age 3 | 27.0 | 14.9 | 35.0 | 30.9 | 30.8 | 23.3 | 19.2 | 16.9 | 17.3 |
| Fishable stock | 2170 | 1916 | 2352 | 2019 | 2088 | 1885 | 1584 | 1620 | 1642 |
| Calculated TAC | 1427 | 1200 | 1635 | 1265 | 1420 | 1285 | 975 | 1050 | 1040 |
| Advised TAC | 850 | 1390 | 1600 | 1265 | 1200 | 1000 | $* * 1090$ | 1350 | 1000 |

*In January 1993, 80000 t were added to the 820000 t recommended after the October 1992 survey due to an unexpected large increase in mean weights.
** In March 2001, 100000 t were added to the 990000 t recommended after the January/February 2001 survey due to much higher mean weights in the catch during 1 February-10 March than measured during the survey.

Table 5.7.1 Capelin in the Iceland-East Greenland-Jan Mayen area. Recruitment of 1-year-old fish (unit $10^{9}$ ) and total stock biomass (' 000 t ) are given for 1 August. Spawning stock biomass (' 000 t ) is given at the time of spawning (March next year). Landings (' 000 t ) are the sum of the total landings in the season starting in the summer/autumn of the year indicated and ending in March of the following year.

| Year | Recruitment | Total <br> Stock biomass | Landings | Spawning <br> stock biomass |
| :--- | ---: | ---: | ---: | ---: |
| 1978 | 164 | 2832 | 1195 | 600 |
| 1979 | 60 | 2135 | 980 | 300 |
| 1980 | 66 | 1130 | 684 | 170 |
| 1981 | 49 | 1038 | 626 | 140 |
| 1982 | 146 | 1020 | 0 | 260 |
| 1983 | 124 | 2070 | 573 | 440 |
| 1984 | 251 | 2427 | 897 | 460 |
| 1985 | 99 | 2811 | 1312 | 460 |
| 1986 | 156 | 3106 | 1333 | 420 |
| 1987 | 144 | 2639 | 1116 | 400 |
| 1988 | 81 | 2101 | 1037 | 440 |
| 1989 | 64 | 1482 | 808 | 115 |
| 1990 | 118 | 1293 | 314 | 330 |
| 1991 | 133 | 1975 | 677 | 475 |
| 1992 | 163 | 2058 | 788 | 499 |
| 1993 | 144 | 2287 | 1179 | 460 |
| 1994 | 224 | 2287 | 864 | 420 |
| 1995 | 197 | 3007 | 929 | 830 |
| 1996 | 191 | 2885 | 1571 | 430 |
| 1997 | 165 | 2348 | 1245 | 492 |
| 1998 | 168 | 2197 | 1100 | 500 |
| 1999 | 138 | 2315 | 933 | 650 |
| 2000 | 140 | 2114 | 1071 | 450 |
| 2001 | 110 | 2439 | 1249 | 475 |
| 2002 | $* 96$ | $* 1770$ | 988 | 410 |

[^3]Figure 5.5.2.1. The relationship between the measured numbers of immature 1 -group capelin in autumn acoustic surveys and the numbers of maturing capelin in 1 August of the following year (left hand figure) and between measured total numbers of 2-group capelin and the maturing 3-group capelin in the following year (right hand figure).


Figure 5.5.2.2. The relationship between the total numbers in the maturing stock and the mean weight of maturing 2group (left hand figure) and 3-group (right hand figure) capelin in autumn 1989-2001.


## BLUE WHITING

### 6.1 Stock Identity and Stock Separation

Blue whiting stock is treated as a single stock for assessment purposes although morphological, physiological, and genetic research has suggested that there may be several components of the stock which mix in the spawning area west of the British Isles (ICES 2000/ACFM:16; Heino et al. WD b).

### 6.1.1 ACFM advice and management applicable to 2002 and 2003

In 1998 ACFM defined limit and precautionary reference points for this stock: $\mathbf{B}_{\text {lim }}\left(1.5\right.$ mill.t) $, \mathbf{B}_{\mathrm{pa}}(2.25$ mill.t) $), \mathbf{F}_{\text {lim }}$ ( 0.51 ) and $\mathbf{F}_{\mathrm{pa}}$ (0.32). The advice of ACFM in following years has been given within a framework defined by these reference points.

In 2001 ACFM stated that " the stock is considered to be outside safe biological limits. In recent years the stock has rapidly declined. SSB is estimated to have been at $\mathbf{B}_{\mathrm{pa}}$ in 2000 and will be close to $\mathbf{B}_{\mathrm{lim}}$ in 2001. Fishing mortality has increased from around the proposed $\mathbf{F}_{\mathrm{pa}}$ in 1997, to well above $\mathbf{F}_{\mathrm{pa}}$ in 1998 and 1999, and well above $\mathbf{F}_{\text {lim }}$ in 2000. Total landings in 2000 were 1.4 million $t$, far above the ICES recommended catch of 800000 t . Landings in 2000 mainly consisted of the strong 1996 and 1997 year classes. The strength of incoming year classes is unknown. ICES recommends that the fishery in 2002 for blue whiting in all areas be closed until a rebuilding plan has been implemented".

In 2002 ACFM stated that "the stock is harvested outside safe biological limits. The spawning stock biomass for 2001 at the spawning time (April) is inside safe biological limits while the SSB for 2002 is expected to be below $\mathbf{B}_{\mathrm{pa}}$. Fishing mortality has increased rapidly in recent years, and is estimated at 0.82 in 2001. Total landings in 2001 were almost 1.8 million $t$. The incoming year classes seem to be strong. ICES recommends that the fishing mortality be less than $\mathbf{F}_{\mathrm{pa}}$ $=0.32$, corresponding to landings of less than 600000 t in 2003". Implementation of a rebuilding plan is not necessary since according to the current assessment the state of the stock is better than previously estimated.

In December 2002 EU, Faroe Islands, Iceland, and Norway agreed to implement a long-term management plan for the fisheries of the blue whiting stock, which is consistent with a precautionary approach, aimed at constraining the harvest within safe biological limits and designed to provide for sustainable fisheries and a greater potential yield. The plan shall consist of the following:

1. Every effort shall be made to prevent the stock from falling below the minimum level of Spawning Stock Biomass (SSB) of 1500000 tonnes.
2. For 2003 and subsequent years, the Parties agreed to restrict their fishing on the basis of a TAC consistent with a fishing mortality less than 0.32 for appropriate age groups as defined by ICES, unless future scientific advice requires modification of the fishing mortality rate.
3. Should the SSB fall below a reference point of 2250000 tonnes ( $\boldsymbol{B}_{p a}$ ) the fishing mortality rate, referred to under paragraph 1, shall be adapted in the light of scientific estimates of the conditions then prevailing. Such adaptation shall ensure a safe and rapid recovery of the SSB to a level in excess of 2250000 tonnes.
4. In order to enhance the potential yield, the Parties shall implement appropriate measures, which will reduce catches of juvenile blue whiting.
5. The Parties shall, as appropriate, review and revise these management measures and strategies on the basis of any new advice provided by ICES.

The agreed management plan is, however, not implemented. In the absence of agreements on a TAC for 2002 and 2003 and the allocation of the TAC, the Coastal States and the Russian Federation implemented unilateral measures to limit blue whiting catches for these years.

### 6.2 Fisheries in 2002

Total catch figures in 2002 were provided by members of the WG. They were estimated to be 1554995 t compared to 1780170 t last year. Time-series with catches by nations and area are given in Tables 6.2.1-6.2.7.

Spatial and temporal distribution of the catches of blue whiting in 2002 is given by quarter and ICES rectangles in Figure 6.2.1. The distribution of the catch by ICES rectangles for the whole year is given in Figure 6.2.2. In 2002 the catch provided as catch by rectangle represented approximately 1.52 million t ( $97.7 \%$ ).

### 6.2.1 Description of the national fisheries

## Denmark:

The Danish blue whiting fishery was conducted by trawlers using a minimum mesh size of 40 mm in the directed fishery and in the fisheries where blue whiting was taken as bycatch, trawls with mesh sizes between 16 and 36 mm were used. The directed fishery caught 39100 t mainly in Divisions IIa (13 600 t ), IVa (20 900 t ) with small catches from Divisions IIIa, Vb, VIa and VIIb. Bycatches of blue whiting (12 100 t ) were caught mainly in the Norway pout fishery in the North Sea and in the Skagerrak. Some blue whiting bycatches were also taken during the human consumption herring fishery in the Skagerrak.

## Germany:

No information available.

## Faroe Islands:

In 2002 Faroes had no agreement with EU on blue whiting and consequently no fishery was conducted in EU waters in 2002. In January the fishery (8 combined purse seiners/trawlers) concentrated on the western and south-western part of the Faroese EEZ (ICES Division Vb). In February catches were still taken in Vb and also in the south-western part of Division VIb outside the EU zone. In March the fishery continued in VIb and XII outside the EU zone. In April catches were taken in the northern part of VIb and in the southern part of the Faroese zone $(\mathrm{Vb})$. In May the fishery continued in the southern area and west off the Faroe plateau $(\mathrm{Vb})$, indicating that the fish migrated west of the Faroes on their way north. The fishery moved northwest towards the Icelandic border and into the Icelandic zone (Divisions Va and Vb ) in late May. In June the fishery operated in Divisions IIa and Va north of $63^{\circ} \mathrm{N}$ (Faroe and Iceland zones) and continued in this area, including Division Vb in July and August. The fishery continued on the Faroe-Iceland ridge (Divisions Va, Vb and IIa) throughout the rest of 2002 (i.e. September-December), gradually moving closer towards the Faroe Islands at the end of the year. All catches were taken with pelagic trawl ( 44 mm mesh size in the codend). The industrial fleet ( 3 trawlers) operated mainly in Norwegian waters (ICES Division IVa) in 2002 with some catches of blue whiting scattered throughout the year.

## France:

No information available.

## Iceland:

Iceland set a total blue whiting quota of 283000 t for the Icelandic and Faroese EEZs and international waters for 2002. Iceland and Faroes have a bilateral agreement of mutual fishing rights for blue whiting within each other's EEZs.

A total of 19 vessels participated in the directed fishery, which started in March in international waters west of the British Isles (ICES Divisions XII, VIb) and small catches in Icelandic waters at SE-Iceland. All the catches were taken using mid-water trawls with a mesh size in the codend of 40 mm . In April the main fishery had moved further north and was largely conducted in Faroese waters SW of the Faroes, but with some catches in Icelandic waters.

In early May the fishery was mainly conducted in the Faroese zone but shifted gradually into the Icelandic zone in the latter half of May. In June part of the fleet was engaged in the fishery for NSSPH in the Norwegian Sea, but most of the blue whiting catch of 22000 t was fished in the Icelandic area. From July to September most of the catches were taken in the Icelandic zone with a gradual increase of catches in the Faroese zone. In October about 13000 t of a total catch of 21000 t were taken in the Faroese zone. In November the fishery came to a halt because the quota set for the Icelandic fleet was reached. The small catch taken in November was mainly from the Icelandic EEZ. The total Icelandic catch was 286540 t .

Iceland has set size limitations on landings of blue whiting. If the catch consists of $30 \%$ or more fish smaller than 25 cm , a temporary area closure is imposed.

## Ireland:

The Irish fishery for blue whiting developed in response to severely restricted quotas for mackerel and herring in the 1990s. Catches peaked in 1998 with approximately 46000 t landed. Since then the imposition of an EU TAC and the allocation of a low quota to Ireland have caused the fishery to contract. In 1998 seven Irish vessels were involved in the fishery. Six vessels fished the small quota of 17165 t allocated to Ireland in 2003. Fishing takes place in February and March between the Porcupine and Rockall after the completion of the spring mackerel fishery.

The fishery is prosecuted by Refrigerated Sea Water trawlers fishing large single trawls that have been specially modified to take bulk catches from deep water. Circumference of the gear may be as great as 1700 m with a brailler mesh of 35 to 40 mm . While the vast majority of Irish landings of blue whiting are for fishmeal production, trials in 1997 and 1998 showed that fish of human consumption standard could be landed. In 2001 approximately 500 t were block frozen by Irish fish processors.

## Netherlands:

Fishing directed to blue whiting takes place by Dutch pelagic trawlers mainly in areas VIa and VIIc in the $1^{\text {st }}$ and $2^{\text {nd }}$ quarter of the year using mesh size of 40 mm . The total catch in 2002 was restricted by a share in a TAC set by the EU. All catches are landed frozen for human consumption. The total quota for the Netherlands in 2002 was 27044 t .

## Norway:

Norway set a blue whiting quota of 250000 t for the Norwegian EEZ, Jan Mayen zone and international waters for 2002. In addition, through international agreements, 120000 t in the EEZ of EU and 35000 t in the Faroese zone were made available to the Norwegian fishery. The mixed industrial fishery in the North Sea/southern Norwegian Sea was allowed to take 79396 t. The total quota for Norwegian vessels in 2002 was 484396 t .

In June-August 2002 an experimental fishery, regulated by specific vessel quotas, was arranged in the Norwegian zone north of $65^{\circ} \mathrm{N}$. Participating vessels were obliged to report certain biological data to the Institute of Marine Research. The total catch from this fishery was about 36000 t , mostly from the area immediately north of $65^{\circ} \mathrm{N}$.

The main Norwegian fishery is a directed pelagic trawl fishery, regulated by vessel quotas, and is carried out on and west of the spawning areas west of the British Isles. The Norwegian fishery in 2002 started at the beginning of February in international waters off Porcupine Bank, then moved northward toward the Rockall area. At the end of March/beginning of April the main fishery took place off the Hebrides area. From there the fishery moved into Faroese waters. The Norwegian fishery in the spawning area was stopped on 5 May when the quota in the EU zone was taken.

Young blue whiting were fished in the North Sea and in the southern Norwegian Sea (areas south of $64^{\circ} \mathrm{N}$ ) in the mixed industrial fishery targeting blue whiting and Norway pout. An estimated catch of approximately 98000 t was taken in this fishery in 2002. More information on this fishery is in Section 1.5 (see also Heino et al., WD a).

## Portugal:

In the Portuguese fisheries, blue whiting is a bycatch in the trawl fishery. Most of the landings come from bottom fish trawlers, but a small percentage also comes from crustacean trawlers. Artisanal fisheries (small purse seines and gillnets) reported a small percentage (less than $5 \%$ of the total Portuguese landings) in the last 5 years.

## Russia:

The Russian blue whiting fishery in 2002 continued from January to December in different NEA areas. In January and February fishing took place mostly in the Faroese EEZ (Vb1). Further, following spawning migrations, the fishing fleet displaced southwards and operated in international waters to the west of the British Isles (XII) until the middle of April. At the end of April, following blue whiting feeding migrations, Russian fishing vessels moved to the Faroese and Norwegian EEZs and international waters in the Norwegian Sea ( $\mathrm{Vb} 1, \mathrm{IIa}$ ) and fished there till the end of September. From October to December a Russian fleet operated mostly in international waters and in the Faroese zone (IIa, Vb1).

The directed blue whiting fishery occurred in all seasons by large fishing vessels using trawls with mesh sizes of 35 to 40 mm .

## Spain:

The Spanish blue whiting fishery is carried out mainly by bottom pair trawlers in a directed fishery and by single bottom trawlers in a bycatch fishery, both using a minimum mesh size of approximately 55 mm . The pair trawler fleet landed 16352 t , taken mainly on the border between Divisions VIIIc and IXa. Bycatches of blue whiting ( 1134 t ) were caught mainly in the bottom trawl fishery in Spanish waters in Divisions VIIIc and IXa, small quantities ( 20 t) were also caught by longliners. These coastal fisheries have trip durations of 1 or 2 days and catches are for human consumption. Thus, coastal landings are rather stable due mainly to market forces.

## Sweden:

No information available.

## UK (Scotland):

No information available.

### 6.3 Biological Characteristics

### 6.3.1 Length composition of catches

Data on the combined length composition of the 2002 commercial catch by quarter of the year from the directed fisheries in the Norwegian Sea and from the stock's main spawning area were provided by the Faroes, Iceland, Ireland, the Netherlands, Norway, Scotland and Russia. Length composition of blue whiting varied from 13 to 42 cm , with $90 \%$ of fish ranging from 21 to 32 cm in length with a mean of 26.4 cm (Table 6.3.1.1). Length compositions of the blue whiting catch and bycatch from "other fisheries" in the Norwegian Sea (south of $64^{\circ} \mathrm{N}$ ), the North Sea and Skagerrak were presented by Norway and Denmark (Table 6.3.1.2). The catches of blue whiting from the mixed industrial fisheries consisted of fish with lengths of 13 to 36 cm and a mean of 22.7 cm . Spain and Portugal caught blue whiting in the Southern area. The Spanish data used for length distribution of catches showed a length range from 15 to 29 cm with a mean length of 20.8 cm (Table 6.3.1.3).

### 6.3.2 Age composition of catches

For the directed fisheries in the northern area in 2002, age compositions were provided by the Faroe Islands, Iceland, Ireland, Norway, the Netherlands and Russia, and the sampled catch accounted for $91 \%$ of the total catch. Estimates of catch in numbers for unsampled catches were raised according to the knowledge of how, where, and when the catches were taken. The age compositions in the directed fisheries are given in Table 6.3.2.1.

Age compositions for blue whiting bycatches from "other fisheries" in the North Sea and Skagerrak were provided by Norway and sampled catch accounted for $78 \%$ of catches. These data were used for allocation of the remaining part of the total in that area. The age compositions are given in Table 6.3.2.2.

For the fisheries in the Southern area, age compositions representing $91 \%$ of the catch were presented by Spain. The age compositions in the southern fishery are given in Table 6.3.2.3.

The combined age composition for the directed fisheries in the Northern area, i.e. the spawning area and the Norwegian Sea, as well as for the bycatch of blue whiting in "other fisheries" and for landings in the Southern area, were assumed to represent the overall age composition of the total landings for the blue whiting stock. The catch numbers-at-age used in the stock assessment are given in Table 6.3.2.4. The 1999 and 2000 year classes were the most numerous in the catches, followed by the 1998, 1997 and 2001 year classes. To calculate the total international catch-at-age, and to document how it was done, the program SALLOC was used (ICES 1998/ACFM:18). The allocations are shown in Annex I.

### 6.3.3 Weight-at-age

Mean weight-at-age data were available from the Faroes, Iceland, Ireland, the Netherlands, Norway, Russia, and Spain. Mean weight-at-age for other countries was based on the allocations shown in Annex I ("ALLOC" files) and was estimated by the SALLOC program for the total international catch. Table 6.3.3.1 shows the mean weight-at-age for the
total catch during 1981-2002 used in the stock assessment. The weight-at-age for the stock was assumed to be the same as the weight-at-age for the catch.

### 6.3.4 Maturity-at-age

Maturity-at-age used in the assessment was obtained by combining maturity ogives from the southern and northern areas, weighted by catch in numbers-at-age (ICES 1995/Assess:7). These are the same as those used since 1994 (Table 6.3.4.1). Although the values of maturity-at-age probably are too low, sufficient information for estimating new ogives is not available.

### 6.3.5 The value of the natural mortality $M$ for blue whiting

The possible need for revising the current estimate of instantaneous natural mortality rate $M$ for blue whiting was discussed in detail by the 2002 WG . Although it was admitted that the current estimate $M=0.2 \mathrm{yr}^{-1}$ might be too low, the factual basis for revision was ambiguous. More recent methodological work by WGMG (ICES 2003/D:03) emphasises that natural mortality rate cannot be estimated reliably with information normally available for stock assessment models. WG therefore considers that there is no new information that would justify a revision of the current estimate of M.

### 6.4 Stock estimates

### 6.4.1 Acoustic surveys

The time-series from the acoustic surveys are given in Tables 6.4.1.1-6.4.1.3.

### 6.4.1.1 Surveys in the spawning season

The Norwegian research vessel R.V. "Johan Hjort" surveyed the blue whiting stock on the shelf edge and bank areas west of the British Isles and in the southern Faroese waters from 29 March - 27 April (Heino et al., WD c). The highest abundance of blue whiting was observed along the shelf edge west from the Hebrides, on the north-western parts of Rockall Bank, and on banks south of the Faroe Islands. The densest aggregations were observed west and northwest of the Hebrides and south of the Faroe Islands (Figure 6.4.1.1.1). The overall distribution of blue whiting was more northern than usual, probably because of the late timing of the survey. Almost all sampled fish were either postspawners or immatures. The spawning stock was estimated to be 10.4 million t ( $132 \times 10^{9}$ individuals), while the immature part of the stock in the survey area was estimated at 1.0 million $\mathrm{t}\left(28 \times 10^{9}\right.$ individuals). Both the numbers and biomass of the spawning stock were marginally lower than the estimate of 2002. The age-stratified estimate of the total stock for 2003 is given in Table 6.4.1.1.1 The spawning stock continues to be dominated by two very strong year classes, 1999 and 2000 (Figure 6.4.1.1.2), which account for about $75 \%$ of the spawning stock biomass. Relatively stable spawning stock biomass between 2002 and 2003 is probably due to the recruitment of more individuals from the abundant 2000 year class to the spawning stock. Immature blue whiting (mostly of age 1) occurred mostly in the southern and northern parts of the survey area, on Porcupine Bank and in the Faroes/Shetland area. Year class 2001 appears to be much less numerous than the strong 2000 year class. The decreasing trend in stock size observed between 1999 and 2001 has now reversed mainly due to the presence of two very strong year classes (1999 and 2000) in the spawning stock the last two years. Examination of the development of year class abundance over consecutive years suggests that catchabilities in the 2002 and 2003 surveys were higher than in earlier surveys, whereas in 2001 the catchability was lower than normal.

Russian investigations in the spawning area were carried out by research vessels "Atlantniro" and "Smolensk" from 11 March to 22 April, (Oganin et al. WD) (Table 6.4.1.1.2). The area from $49^{\circ} 50^{\prime} \mathrm{N}$ to $61^{\circ} 30^{\prime} \mathrm{N}$ inside the EU EEZ and from $53^{\circ} 00^{\prime} \mathrm{N}$ to $59^{\circ} 52^{\prime} \mathrm{N}$ in the international waters was surveyed. The spawning stock biomass estimate was 16.6 million t , whereas the immature part of the stock was 1.9 million t . The densest concentrations were observed in the northwestern slope of the Porcupine and extended to the north between $13^{\circ} 00^{\prime}-14^{\circ} 00^{\prime} \mathrm{W}$. The spawning stock was dominated by the 1998 to 2000 year classes.

The three research vessels, R.V. "Atlantniro" and R.V. "Smolensk" from Russia and R.V. "Johan Hjort" from Norway, participating in the spawning stock investigations in March-April 2003 deployed similar survey strategies and carried comparable acoustic instrumentation, but the survey tracks and timing of the surveys were different. Norwegian and Russian surveys yield strikingly different abundance estimates in 2003 as was the case in 2002. Consequently the WG decided that a joint estimation of the Norwegian and Russian spawning stock surveys should not be used in the assessment. Furthermore, the estimated age structure is different: the Norwegian estimate is dominated by ages 3 and 4,
whereas the Russian estimates show higher numbers of older fish as well. A number of factors may contribute to these differences: 1) temporal and spatial coverage of the spawning area in relation to migrations of blue whiting, 2) sampling selectivity, with the Norwegian trawl under-sampling large fish relative to the Russian trawls. Estimates of total and spawning biomass of blue whiting in the spawning area made by Russian, Norwegian, and Faroese surveys since 1983 are given in Table 6.4.1.1.3. Usually, acoustic estimates have been well above the analytical assessments. Therefore the acoustic estimates have been used only as relative indices. A factor contributing to the high acoustic estimates is the use of a target strength that is probably too low (Heino et al. WDc).

### 6.4.1.2 Surveys in the feeding area

Since 1995, Norway, Russia, Iceland, and Faroes, and in the period 1997-2000 the EU, have coordinated their survey effort on pelagic fish stocks in the Norwegian Sea. ICES 2002/D:07 reported on distributions and migrations of blue whiting in 2002.

A Russian survey in June covered the northeastern Icelandic EEZ, the Jan Mayen zone and central Norwegian Sea. Blue whiting was observed in most of the area with the main concentrations in the east between $0^{\circ}$ and $07^{\circ} \mathrm{E}$. Another area with relatively high densities was observed in international waters between Jan Mayen and Lofoten. Echo-recordings of blue whiting were usually registered as a scattered layer at depths of $150-300 \mathrm{~m}$ and the lengths ranged between 19-36 cm , with fish of 21-28 cm dominating the length distribution. An age-disaggregated estimate is not available but the length distribution suggests the 2000 year class is dominant. The total biomass estimated in June was 1.8 million t and an abundance of $20.7 \times 10^{9}$ fish (ICES 2002/D:07).

Another Russian survey in July covered the central and northern Norwegian Sea. The main recordings of blue whiting were observed in an area between $0^{\circ}$ and $10^{\circ} \mathrm{E}$ in the southeastern part of the survey area. The biomass estimate in the survey area was 1.1 million t , consisting of $13.8 \times 10^{9}$ individuals. The length ranged between $19-31 \mathrm{~cm}$ with fish of 2227 cm dominating the length distribution. An age-disaggregated estimate is not available but the length distribution points to the dominance of the 2000 year class (ICES 2002/D:07).

The Norwegian and Icelandic surveys conducted in July-August from 1998 to 2001 and the Icelandic survey in 2002 covered approximately the same area from year to year. There was no comparable Norwegian survey in 2002, but a Norwegian and Faroese survey in May covered similar areas. For both the Icelandic EEZ and Norwegian Sea, agestratified estimates of blue whiting were reported. These are given in Tables 6.4.1.2.1 and 6.4.1.3, respectively.

There was a steady downward trend in the biomass estimate in the surveys from 1998 to 2000. The biomass estimated in 2001 and 2002 was more than two times higher than the 2000 estimate, as shown in the text table below.

| Year | Norwegian survey <br> $($ million t) | Icelandic survey <br> $($ million t) | Total <br> $($ million t) |
| :--- | :--- | :--- | :--- |
| 1998 | 6.6 | 1.6 | 8.2 |
| 1999 | 4.2 | 1.8 | 6.0 |
| 2000 | 2.5 | 1.2 | 3.7 |
| 2001 | 5.9 | 2.1 | 8.0 |
| 2002 | $6.0^{*}$ | 1.9 | 7.9 |
| $*$ |  |  |  |

Biomass estimates of blue whiting from the Faroese area and the Norwegian Sea in May 2002 were similar to that of the 2001 survey in July. The 2000 year class was still very prominent ( $56 \%$ in number) in the survey. This confirms the previous belief that 2000 is a large year class. Currently the blue whiting stock is dominated by the 1999 and 2000 yearclasses. The biomass estimate of blue whiting in the Icelandic area was also much higher in 2001 and 2002 than in 2000 ( $67 \%$ and $51 \%$ higher). The 2001 year class was very numerous in the Icelandic survey as 0 -group and was still quite numerous in the 2002 survey ( $46 \%$ in number).

### 6.4.2 Bottom trawl surveys in the southern area

Bottom trawl surveys have been conducted off the Galician (NW Spain) and Portuguese coasts since 1980 and 1979 respectively, following a stratified random sampling design and covering depths down to 500 m . Since 1983, the area
covered in the Spanish survey was extended to completely cover Spanish waters in Division VIIIc. The area covered in the Portuguese survey was also extended in 1989 to the $750-\mathrm{m}$ contour. A new stratification in the Spanish surveys has been established since 1997. Stratified mean catches and standard errors from the Spanish and Portuguese surveys are shown in Tables 6.4.2.1 and 6.4.2.2. In both areas larger mean catch rates are observed in the $100-500 \mathrm{~m}$ depth range. Since 1988 the highest catch rates in the Spanish survey were observed in 1999 ( $124 \mathrm{~kg} / \mathrm{haul}$ ). The 2001 estimate is relatively low in both the Spanish ( $52 \mathrm{~kg} /$ haul) and the Portuguese autumn survey ( $116 \mathrm{~kg} / \mathrm{haul}$ ). The Portuguese autumn surveys generally give higher values than in the summer surveys, and a better correlation with the Spanish surveys (Figure 6.4.2.1).

### 6.4.3 Catch per unit of effort

The Spanish pair trawl CPUE series (Table 6.4.3.1) has been used for several years as a tuning fleet in the blue whiting assessment. This fleet represents only a small part of the landings caught in a small part of the distribution area. Due to this fact, and following a recommendation of the Methods Working Group (ICES 2003/D:03), this tuning series was not used this year in the assessments. The data show a slight decreasing trend in CPUE (Figure 6.4.3.1).

CPUE data from the commercial fleet (pelagic trawl) in the spawning area in 2001-2002 were submitted from Norway and added to the existing time-series. The CPUE is calculated as an overall aggregated CPUE in the Norwegian spawning fishery and is based on data from the vessels that have submitted their logbooks (about one-third of the vessels do so). The data suggest an overall increasing trend in CPUE since early 1990s (Figure 6.4.3.2). However, no attempt has been made to correct the data for the increase in catchability caused by technological improvements. As in previous years, the data have not been used in the assessment.

### 6.4.4 Data exploration

At the meeting of the Methods WG in January 2003, five different assessment models have been applied on blue whiting data (AMCI, ISVPA, ICA, XSA and CADAPT). The evaluation of these models partly addressed a request to ICES to evaluate some of these models with respect to blue whiting. The differences in results from the different models have been discussed by that group and are compared in Figure 6.4.4.1. The Methods WG gave the following recommendations to WGNPBW:

- The choice of appropriate model to assess the stock is not clear-cut and the approach used at this meeting of WGMG of exploring a number of competing models is to be commended as an aid to disentangle the apparent conflicting sources of data.
- To exclude the Spanish CPUE series from the calibration process as this survey was clearly shown to have no relevant information to the overall stock.
- To exclude the acoustic series which end in the 1980s because these surveys only provide information on cohorts that have been fished out by now. Alternatively it could be considered whether the two survey periods could be combined if a gear effect can be estimated for the introduction of the new EK500 equipment.
- It was found that the survey data was very noisy and often contradictory. This may be caused by sampling problems in the sampling for age in the acoustic surveys. The WG could explore the possibility of including acoustic survey data as SSB estimates rather than age-disaggregated indices. This would perhaps get rid of the noise in the age signal although the mortality signal would be lost.

WGNPBW addressed some of the recommendations using the same data as WGMG, but including catches in 2002 and the results of the 2003 Norwegian acoustic survey on the spawning grounds.

The assessment models explored by WGNPBW were AMCI and ISVPA. Different configurations were run with both models with individual and different combinations of tuning series and different settings. In addition a run with ICA was made, using the standard settings.

## Tuning data available to the blue whiting assessments

There are six tuning fleets available for the blue whiting assessment: an acoustic survey in the Norwegian Sea (covering the feeding area of the northern stock component), a Norwegian and a Russian acoustic survey on the spawning grounds, a CPUE series from Spanish pair trawlers, a Spanish bottom trawl survey and a Portuguese bottom trawl survey. The last three fleets cover only a small part of the southern distribution area of the stock. The Spanish and

Portuguese survey fleets have not been used since 1998 and this year the WG decided to leave the CPUE time-series from Spanish pair trawlers out the assessment as well, following a recommendation of the Working Group on Methods on Fish Stock Assessments (ICES 2003/D:03), because this series proved to have no relevant information to the overall stock.

Catch numbers available to the assessment were available from 1981-2002 for ages 0-10+. Discard data were not available to the assessments. Last year 0-group data were included in the assessment. Exploratory runs including the 0group showed the largest residuals in this age group. The catch data of this age group are very noisy, maybe because of changes in the distribution of the fleet for this age group, and provide no information on year-class strength. Therefore they were excluded from the assessment.

### 6.4.4.1 Analysis of catch-at-age in commercial data and in surveys

There has been a remarkable increase in the fishery for blue whiting in recent years. Before 1998, the average annual landings were below 600000 t , while the average for the period 1998 to 2002 is above 1.4 million t . A crucial question is to what extent this large increase in the fishery has been obtained through an increase in the fishing mortality.

## Catch data

The mortality signal in the catches is seen as the decline in catches along the cohorts. Figure 6.4.4.1.1 shows the slopes for log catch-at-age for the year classes from 1990 and earlier, the 91-94 year classes and the 95-97 year classes, and indicates that the recent big ones have been exploited at a $Z$ of about 0.8 , while the earlier ones had a $Z$ at about 0.4. The Z in the early years is lower than we have in most assessments (for ages 5-9).

## Survey data

Figure 6.4.4.1.2 shows the log indices by year class for the Norwegian acoustic survey on the spawning grounds, and calculated slopes of the curves for the average indices at age for the year classes up to 1990, and from the 1991 year class and onwards, and a regression from age 4 to 8 . This indicates mortality in the order of 0.66 for the early period, and 1.1 for the late period. Similar analyses with the Norwegian Sea survey give similar values $(0.55$ and 1.1 respectively, but this is without 2002 data). Thus, the results from these surveys, assuming that the catchabilities at age are the same, indicate a mortality increase as do the catch data, but with higher absolute values.

Altogether, the catch and survey data give quite strong indications that the mortality since about 1998 at ages 4-5 and higher has been higher than previously, although there are some discrepancies with respect to the actual levels. Secondly, the recent survey data have strong year effects, and the estimate of the terminal F is largely a compromise between these conflicting signals, which renders the estimates of the exploitation rate and state of the stock very uncertain.

### 6.4.4.2 Data exploration with AMCI

In order to evaluate robustness of the assessment, a set of assessment runs was based on tuning with survey fleets used in the final assessment (Section 6.4.5.1), but using only one fleet at a time. In addition, one run was based solely on catch data. These assessments give consistent results (Figure 6.4.4.2.1), similar to those obtained in the final AMCI assessment. Residuals show some year effects but are similar to those obtained in the final AMCI assessment (Figure 6.4.4.2.2). There are some differences in the absolute levels of the assessment but the patterns are similar. Tuning based on Russian spawning stock survey shows the strongest deviation from the common pattern, indicating higher SSB and lower F during the most recent years. However, this time-series ends in 1996 and cannot therefore provide much information on the stock in later years.

Recruitment in 2001 (year class 2000) could not be estimated reliably. Exploratory AMCI runs where estimation was attempted did suggest strong recruitment at a level similar to recruitment in 2000, but the estimate was quite variable and, moreover, caused instability in other model parameters. Information from the surveys in 2002 and 2003 as well as from the PGSPFN in 2001 and 2002 (Holst et al. 2001, ICES 2002/D:07) suggest that year class 2000 is at least as strong as year class 1999. Recruitment in 2001 was therefore set to be similar to recruitment in 2000, $30 \times 10^{9}$ individuals at age 1. Recruitment in 2002 was set to a value similar to geometric mean recruitment in the period 1981-2000, $11.5 \times 10^{9}$.

The Methods Working Group recommended considering treating data from the spawning stock surveys as SSB indices. This was attempted with the Norwegian time-series on the spawning ground, which is longer than the Russian time-
series. The results are presented in Figure 6.4.4.2.3. The resulting assessment shows much higher SSB in the recent years than any other AMCI assessment carried out by the WG. In addition, the model suggested a decreasing trend in fishing mortality, which is at odds with the patterns indicated by log catch ratios. The Working Group acknowledged that tuning with only SSB data was not reliable because of the known year effects in this time-series. Therefore the AMCI run with the three acoustic fleets (age structure) was adopted as the final AMCI run.

In previous years, CPUE time-series from Spanish pair trawlers have been included in the blue whiting assessments. This data suggest somewhat different dynamics than the blue whiting stock as the whole, and the Working Group on Methods on Fish Stock Assessments (ICES 2003/D:03) recommended exclusion of this time-series from the tuning data. Following this recommendation, the WG chose not to use the Spanish CPUE data in the final assessment. For comparison reasons a run is presented which included the Spanish data. The results are shown in Figure 6.4.4.2.3. Inclusion of the data in assessment results in lower estimate of SSB and higher estimate of $F$ in the recent years, but it yields no qualitatively different results.

Age-disaggregated abundance estimates of blue whiting in Icelandic waters are available from 1999 onwards (Section 6.4.1.2). These data come from the feeding area of young blue whiting and may thus provide information on recruiting year classes. The data were included in an exploratory run along with final tuning data (Figure 6.4.4.2.3). This data resulted in moderate changes in estimated recruitment in 2000, but the overall effects were very small. The available time-series is so short that its value in providing information on recruitment cannot yet be assessed but is worth further analyses.

### 6.4.4.3 Data exploration with ISVPA

For blue whiting exploratory runs the same version of the ISVPA, which was used in frames of testing at the ICES WG on Methods (ICES 2003/D:03) was used (effort-controlled, unbiased separable representation of fishing mortality coefficients with minimization of the sum of squared errors (SSE) for each component of the ISVPA loss function). The age range for analysis was taken from 1 to $10+$ because of extremely high residuals for age group 0 . Options of the model are listed below.

Implementation of this version of the model gave distinct minima of the respective component of the loss function for each survey and for catch-at-age (Figure 6.4.4.3.1). According to recommendation of the ICES 2003/D:03 the data for CPUE of Spanish fleet was excluded from analysis. The signal from the integral SSB index from Norwegian surveys was also tested and revealed a distinct minimum.

Profiles of total loss function for cases when catch-at-age is used along with all age-structured auxiliary information, as well as with integral SSB index, are shown in Figure 6.4.4.3.2.

Tuning of the model on integral SSB index gives higher estimates of SSB in recent years than tuning on age-structured indexes (Figure 6.4.4.3.3). In the text table below, main settings and options used in ISVPA in WGNPBW 2002, Coastal States Meeting 2002 and WGMG 2003, and WGNPBW 2003 are shown.

| Settings/options for the ISVPA run | 2002(WG) | 2002 (Coastal <br> States meeting) and WG on Methods (2003) | 2003 (WG) |
| :---: | :---: | :---: | :---: |
| Numbers of age-structured tuning series |  | 3 splitted | 3 splitted |
| Version of the model | Effort-controlled | Effort-controlled | Effort-controlled |
| Numbers of biomass tuning series |  | 0 | 0 |
| Constraint | Unbiased separable representation of $F$ | Unbiased separable representation of F | Unbiased separable representation of F |
| Number of years with a separable constraint | 1981-2001 | 1981-2001 | 1981-2002 |
| Reference age for separable constraint | no | no | No |
| Constant selection pattern | yes | Yes | Yes |
| S on the last age | Equal to that for previous age | Equal to that for previous age | Equal to that for previous age |
| Age span for calculation of reference F | 3-7 | 3-7 | 3-7 |
| Weight given to age groups and years in separable period | 1 | 1 | 1 |
| Catchability model for all fleets |  | Constant, q(a) are estimated | Constant, q(a) are estimated |
| Age range for the analysis | 1-10+ | 1-10+ | 1-10+ |
| Survey weights for all ages in all fleets |  | 1 | 1 |
| What is minimized for residuals in logarithmic catch-atage |  | SSE | SSE |
| What is minimized for logarithmic abundance-at-age (for indices) |  | SSE | SSE |

### 6.4.4.4 Data exploration with ICA

In order to explore various interpretations of the assessment input data, various assessment models were used. One run with the model ICA was done, using similar settings as during the 2002 WG (see text table below). The tuning data were changed from 2002. Following the recommendations from the 2003 meeting of the Methods Working Group (ICES CM2003/D:03) one of the tuning fleets, the Spanish CPUE fleet, was left out of the analysis. In addition, the tuning series from the acoustic fleets that previously had been split into two series due to possible effects from changes in acoustic equipment were joined this year, to make the run compatible with those made with AMCI. Accordingly, there were now three age-structured fleets left in the analysis; the Norwegian Spawning survey, the Russian Spawning survey, and the Norwegian Sea acoustic survey. The age range in the analysis was changed from $0-10$ to $1-10$, to be comparable with the final run with AMCI and ISVPA. Main options and settings are given in the text table below, and compared to the settings used in 2002. Some diagnostic plots are given in Figures 6.4.4.4.1-2, and the results are shown in Figure 6.4.4.4.3 and Table 6.4.4.4.1.

| Settings/options for the ICA run | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ |
| :--- | :--- | :--- |
| Numbers of age-structured tuning series | $4^{1}$ | 3 |
| Numbers of biomass tuning series | 0 | 0 |
| Number of years with a separable constraint | 4 | 5 |
| Reference age for separable constraint | 6 | 6 |
| Constant selection pattern | Yes | Yes |
| S to be fixed on the last age | 1.5 | 1.5 |
| Age span for calculation of reference F | $3-7$ | $3-7$ |
| Weight given to age groups and years in separable period | 1 | 1 |
| Catchability model for all fleets | Linear | Linear |
| Age range for the analysis | $0-10$ | $1-10$ |
| Survey weights for all ages in all fleets | varying ${ }^{2}$ | 1 |
| Extent of correlation between errors across ages | 1 | 1 |
| Shrinkage | No | No |

It is seen that the SSQ surface is rather flat for a range of terminal Fs (Figure 6.4.4.4.1) and the minimum is poorly defined. There are negative age residuals for the youngest age groups and positive residuals for the oldest (Figure 6.4.4.4.2). The results show that the stock size has increased considerably in recent years, the spawning stock biomass reaching 4 million t in 2002. However, because of the large catches taken in the period after 1998, the fishing mortality has also increased, and was estimated to be 0.62 in 2002. All year classes after 1994 are estimated to be well above average strength.

### 6.4.5 Stock assessment

### 6.4.5.1 Final assessment with AMCI

The key settings and data for the final blue whiting assessment in 2003 are shown in the table below. The key settings of the final assessment in 2002 are also shown for comparison. Some of the settings are described in more detail after the table.

[^4]| Settings/options for the AMCI run | 2002 | 2003 |
| :---: | :---: | :---: |
| AMCI version | 2.1 | $2.2{ }^{3}$ |
| Age range for the analysis | 0-10+ | 1-10+ |
| Last age a plus-group? | Yes | Yes |
| Age at recruitment (from Jan 1 in the year of spawning) | 0.5 | 1 |
| Recruitment in the terminal year | Fixed | Fixed |
| Recruitment in the terminal year-1 | Estimated | Fixed |
| Catch data |  |  |
| Weights for the partial objective functions for the catch fleet |  |  |
| Log sum of squares of catches-at-age | 1 | 1 |
| Log sum of squares of yearly yields | 1 | 1 |
| Weights of catch-at-age, age 0 and 1 years | 0.1, 0.5 | n.a., 0.5 |
| Constant selection pattern for the catch fleet? | Almost | Almost |
| Selectivity for age 10 equals average of selectivity at age 8-9? | No | Yes |
| Age-structured tuning time-series |  |  |
| Norwegian acoustic survey on the spawning grounds, ages 2-8 | 1981-2002 | 1981-2003 |
| Flat selectivity for ages 6-8? | Yes | No |
| Weight in tuning for the partial objective function | 1 | 1 |
| Russian acoustic survey on the spawning grounds, ages 3-8, | 1982-1996 | 1982-1996 |
| Flat selectivity for ages 7-8? | Yes | No |
| Weight in tuning for the partial objective function | 1 | 1 |
| Norwegian Sea acoustic survey, ages 1-7 | 1981-2001 | 1981-2001 |
| Flat selectivity for ages 5-7? | Yes | No |
| Weight in tuning for the partial objective function | 1 | 1 |
| CPUE time-series from Spanish pair trawlers, ages 1-6 ${ }^{4}$ | 1983-2001 | not used |
| Flat selectivity for ages 5-6? | Yes | n.a. |
| Weight in tuning for the partial objective function | 1 | n.a. |
| Piece-wise constant selection pattern for the tuning fleets? | Yes | Yes |
| Biomass tuning time-series | 0 | 0 |

Survey data used in tuning are shown in Table 6.4.5.1.1. As in previous years, the three acoustic surveys were split into two time periods reflecting a likely change in catchability caused by a change in acoustic equipment (Simrad EK-500). From 2002 onwards the splitting of these time-series has technically been obtained by estimating age-specific catchability separately for the two periods. Survey indices are treated as relative abundance indices.

Fishing mortality was modelled as separable, but with an allowance for a gradual change in the selection from year to year. The gain factor for change in selection was 0.2 for age 1 , and 0.1 for the older ages. This implies that the selection at age 1 is allowed to vary more according to the year-to-year variation in the catches than the selection at the older ages. The selection at age $10+$ was fixed to the average of ages $8-9$ years.

Recruitment in 2001 was set to $30 \times 10^{9}$ individuals at age 1. Recruitment in 2002 was set to the geometric mean recruitment in the period 1981-2000, $11.5 \times 10^{9}$.

[^5]Catch-at-age data are input at yearly resolution (Table 6.3.2.4). However, AMCI operates internally on a quarterly basis. The spawning stock is derived from the mean stock numbers in the first quarter, and the survey indices are related to the mean values in the survey season (Table 6.4.5.1.1.). The yearly fishing mortality was split on quarters assuming that the proportion 0.35 of the total annual fishing mortality occurs in the first and in the second quarter, 0.2 in the third quarter, and 0.1 in the fourth quarter.

The model was run until 2005. The results for 2003 and onwards, except the SSB in 2003, are predicted values assuming the same fishing mortality as in 2002 and constant recruitment. The key results are presented in Tables 6.4.5.1.2-6.4.5.1.4 and summarized in Figure 6.4.5.1.1. Residuals of the model fit are shown in Figure 6.4.5.1.2. Some cohort effects are visible in the catch residuals for the early cohorts, and year effects occur throughout the survey timeseries. The minimum SSQ (between the fitted model and the observed data) is not well defined. A range of terminal F's can give SSQ that are close to the global minimum (Figure 6.4.5.1.3). Thus, the data do not allow unique characterisation of the stock in the most recent years. Selection pattern in terms of age- and year-specific F's is shown in Figure 6.4.5.1.4.

The assessment (Table 6.4.5.1.4) indicates that fishing mortality has increased sharply in recent years. The exploitation pattern has been relatively stable with the major exploitation being on adults. SSB has increased compared to the period from the early 1980s to the late 1990s. However, it has been declining since 1999 and is expected to continue doing so if fishing mortality remains at $\mathrm{F}=0.48 \mathrm{yr}^{-1}$ estimated for 2002 and assumed recruitment. Recruitment of the 1996 year class is the highest in the time-series. All following year classes are also large, indicating a possible change in recruitment dynamics of the stock. Even the weakest year class born after 1996, that of 1998, is much stronger than was typical in the previous period.

A bootstrap run (Figure 6.4.5.1.5) gives an indication of the uncertainty in the assessment. Even though the bootstrap replicates reproduce similar temporal patterns in recruitment, spawning stock biomass and fishing mortality as the final assessment, uncertainty in the absolute level of these metrics during the recent years is clearly visible. Moreover, the development of SSB and fishing mortality after 1999 remain highly uncertain.

Retrospective analysis (Figure 6.4.5.1.6) shows that the assessments with 2001 or 2002 as the terminal year are very consistent. However, 1999 or 2000 as the terminal year yield assessments suggesting lower SSB and recruitment and higher F .

### 6.4.5.2 Final assessment with ISVPA

For the final run the model was run with the same settings as for exploratory runs, but only signals from age-structured auxiliary information were used in addition to signals from catch-at-age. The results of the stock assessment are given in the Tables 6.4.5.2.1-6.4.5.2.3.

Comparison of theoretical catches and reported ones for age groups, included into analysis, are shown in Figure 6.4.5.2.1.

Residuals for catch-at-age and each survey data (in the minimum of the ISVPA loss function) are shown in Figure 6.4.5.2.2. ISVPA residuals in logarithmic catch-at-age residuals have apparent cohort structure. This structure may be dealt with by a large period of separability constraint, used to increase the stability of estimates of selection factors. The cohort peculiarities in the matrix of residuals may be considered as a negative feature, but, as it was mentioned in the Report of the Working Group on Methods (ICES 2003/D:03), the 1992 year class has a strong anomaly in its catch dynamics for young ages and it may not be unreasonable to retain high residuals for this year class, which means that the estimates of the model parameters are less based on this cohort. The structure of residuals for indices is less certain and includes some elements of both cohort and year effects.

The selection pattern in terms of age- and year-specific F's is shown in Figure 6.4.5.2.3.
In the retrospective analysis the same settings were retained even though some of the indices might become uninformative because of the shortening of the year range. This may be the reason for the instability of the results for 1998-2000. The results of the retrospective analysis are shown in Figure 6.4.5.2.4.

Results of the bootstrap estimation of confidence intervals are shown in Figure 6.4.5.2.5. For effort-controlled version of the model, used for blue whiting, the procedure consists in application of conditional parametric bootstrap (assuming lognormal distribution of residuals) with variance, estimated in the basic run; a lognormally-distributed random noise with variance $=0.3$ for each survey was added to the age-structured indices.

### 6.4.5.3 Comparison between AMCI and ISVPA assessment

The results of the assessments carried out with the AMCI and ISVPA model are compared in Figure 6.4.5.3.1.

SSB: The models show reasonable agreement in the historical estimates of SSB except for the most recent years. Both models indicate a significant increase in SSB in the late 1990's to about 4 million $t$. In the period thereafter, SSB continues to increase to a historic high (about 6 million $t$ ) in 2002 in the ISVPA assessment but has decreased slightly in the AMCI assessment to about 3.5 million t .

Fishing mortality: Trends in fishing mortality estimated by both models are similar in the historical period. However, historical ISVPA estimates are consistently higher than those by AMCI. Both models indicate a sharp decrease in F from 1990 to 1991. The reasons for this decrease are unclear and may reflect a shift in the fishery to other components of the stock. Also, both models indicate an increase in F after 1994, which continues to increase to a historic high in 2002 in the AMCI assessment. However, ISVPA shows a decrease in F after 1998.

Recruitment: Both models indicate that recruitment has increased after 1995. The 1996 year class was estimated as the strongest in the time-series. The estimate by both models is almost the same. The estimates of more recent year classes by ISVPA are consistently higher than those by AMCI.

The models give a different interpretation of the large increase in catches by the fishery since 1998. ISVPA explains the increase of the catches almost entirely by the increase in biomass resulting from the large recruitment. The model supports this interpretation unambiguously, in a sense that the minimum SSE is well defined (Figure 6.4.4.3.2). AMCI, on the other hand, explains the increase of the catches partly by a large increase in fishing mortality and partly by an increase in biomass. However, the SSE surface around the minimum is rather flat (Figure 6.4.5.1.3), indicating that solutions with somewhat higher or lower F would fit the data almost as well.

In the absence of quantitative data demonstrating changes in the fishery the WG felt it difficult to make a choice between the two models. Nevertheless, the increase in F estimated by AMCI is in line with the exploratory analysis of catch curves and survey data (Section 6.4.4).

### 6.5 Short-term Projections

Based on the final AMCI and ISVPA runs, two deterministic short-term projections were made using the MFDP (version 1a) program. The yield-per-recruit estimations were made by means of the MFYPR (version 2a) program. The input values are shown in Tables 6.5 .1 and 6.5.2, respectively. The weight in the stock and catch were taken from the average of the last three years' values.

In the first short-term projection the selection pattern and the reference F in 2001 from the final AMCI run were used as input values in 2003. The recruitment in 2003-2005 was set as the geometric mean of the recruitment values in the period 1981-2000 in the AMCI run. For all ages the output values in 2002 from the AMCI run were used as the initial stock size. The proportion of F and M before spawning was set to 0.25 , taking into account the proportion of the catches that take place before the spawning period.

The same settings were chosen to run the second short-term projection, based on the values provided by the final ISVPA run. The results are given in Tables 6.5.3 and 6.5.4, and the standard plots are given in Figures 6.5.1 and 6.5.2.

## AMCI

Continued fishing at the 2002 level predicts a catch of 1.3 million t in 2003 and 1.1 million t in 2004. This exploitation rate implies a decreasing trend of SSB with 3.1 million t in 2003 and 2.5 and 2.1 million t in 2004 and 2005, respectively. The predicted total stock biomass will also decrease from 4.9 million t in 2003 to 4.2 and 3.6 million t in the following years.

## ISVPA

Continued fishing at the 2002 level predicts higher catches of 1.6 million $t$ in 2003 and 1.5 million $t$ in 2004. At this exploitation rate SSB is predicted to decrease from 5 million t in 1994 to 4.3 million t in 1995. The ISVP predicted stock biomass for 2003 is 8.5 million $t$ and is predicted to fall to 7.5 million and 6.5 million $t$ over the following two years.

The Working Group did not carry out medium-term projections. In general, medium-term projections in the most recent period are largely defined by the presently assumed stock size, recruitment and exploitation level, while the period further in the future is mostly affected by assumptions on recruitment dynamics and exploitation level. The Working Group considered that, given the uncertain present situation of the stock and its exploitation, and the poor knowledge on stock dynamics, medium-term projections for blue whiting would be predominantly based on assumptions and are not very instructive (in fact they can be misleading). Also, no standard medium-term projection software was available to the Working Group to carry out the calculations.

### 6.7 Precautionary Reference Points

The present precautionary reference points have been introduced in the advice of ACFM in 1998. The values and their technical basis are:
$\mathbf{B}_{\text {lim }}: 1.5 \mathrm{mill} \mathrm{t} ; \mathbf{B}_{\text {loss }}$
$\mathbf{B}_{\mathrm{pa}}: 2.25$ mill. t; $\mathbf{B}_{\mathrm{lim}} * 1.5$
$\mathbf{F}_{\text {lim }}: 0.51 ; \mathbf{F}_{\text {loss }}$
$\mathbf{F}_{\mathrm{pa}}: 0.32 ; \mathbf{F}_{\text {med }}$.
Although problems have been identified with these reference points they have remained unchanged since 1998. A major problem is that fishing at $\mathbf{F}_{\mathrm{pa}}$ implies a high probability of bringing the stock below $\mathbf{B}_{\mathrm{pa}}$, in other words the present combination of $\mathbf{F}_{\mathrm{pa}}$ and $\mathbf{B}_{\mathrm{pa}}$ is inconsistent.

The present reference points are based on assessments of the stock carried out at the end of the 1990s and are based on a relatively short time-series. Since then major changes have been observed in the stock and in the fishery. This adds to the need to revise the reference points, because of their role as targets for rebuilding and guidelines for future exploitation.

It was the intention by ICES to revisit all precautionary reference points in 2003 for those stocks ICES provides advice for. However, it is not likely that this will be achieved. The SGPRP revisited the $\mathbf{B}_{\text {lim }}$ reference point for blue whiting in February 2003 (ICES 2003/ACFM:15). The current $\mathbf{B}_{\text {lim }}$ value of 1.5 million $t$ was based on an estimate of $\mathbf{B}_{\text {loss }}$ from an assessment in 1998. Since a segmented regression on the stock recruitment data was not significant, $\mathbf{B}_{\text {loss }}$ remains the obvious candidate for $\mathbf{B}_{\text {lim }}$. In the assessment carried out in $2002 \mathbf{B}_{\text {loss }}$ was estimated to be 1.2 million $t$ and SGPRP proposed this value for $\mathbf{B}_{\text {lim }}$.

Based on assessments carried out in 2003 by WGNPBW, the $\mathbf{B}_{\text {loss }}$ is estimated between 1.7 million $t$ (AMCI assessment) and 1.8 million t (ISVPA assessment). The estimates of $\mathbf{B}_{\text {loss }}$ values in recent years vary considerable, however, around the present $\mathbf{B}_{\text {lim }}$. The relatively large changes in the blue whiting assessments and the differences in stock dynamics estimated by different assessment methods make it difficult to estimate a stable value for $\mathbf{B}_{\text {lim }}$. If the value of $\mathbf{B}_{\mathrm{lim}}$ is to be revised, it is likely to be around the region of the present value of 1.5 million tonnes.

The SGPA provided guidelines to re-estimate the other PA reference points based on $\mathbf{B}_{\mathrm{lim}}$ and the historical uncertainty of the assessment. The proposed procedure is quite complicated and requires software still to be developed. The other PA reference points have not been estimated by the WG.

It should be noted that the PA reference points presently applied in the ICES advice are based on an ICA assessment in 1998. Since then regular changes have been made in the assessment by selecting other assessment methods or different assessment configurations.

### 6.8 Management consideration

The fishery for blue whiting has expanded rapidly in recent years, while no agreement on TAC has been reached. The reported catches in 1998 to 2002 were all well above 1 million $t$, reaching 1.7 million $t$ in 2001. The SSB has been at a fairly high level, due to exceptionally good recruitment in recent years. The year classes 1995-1999 are all well above average strength. The estimated strength of the 2000 and 2001 year classes is uncertain, but there is evidence that the

2000 year class is strong. Without the strong year classes recruited in recent years, the intensive fishery would have led to a severe depletion of the stock before now.

However, the assessments made in 2003 give a more optimistic view of the present stock situation compared to that made in 2001 and 2002. The main reason is that the incoming year classes are seemingly stronger than estimated previously. The SSB in 2002 is, according to both alternative assessments made, above the $\mathbf{B}_{\mathrm{pa}}$ presently used.

### 6.8.1 Exploitation of juvenile blue whiting

Although the main fishery for blue whiting takes place in the spawning area where mostly mature fish are caught, some juveniles are caught in all areas. However, the blue whiting catch from the mixed industrial fishery for blue whiting in the North Sea is largely based on juveniles.

The Working Group was able to study age composition in catches from Norwegian mixed industrial fishery in 20002002 (Heino et al., WD a). These data suggest that in most years and areas, the catches are dominated by fish of age 1 year, which are almost entirely juveniles. The Norwegian catch amounts to roughly $500 \cdot 106$ individuals annually. During the period 2000-2001, recruitment at age 1 year has been estimated at about $30 \cdot 109$ individuals. Considering catches from other countries (Table 6.2.3), the catch of age 1 year blue whiting does not represent more than $5 \%$ of the estimated numbers of blue whiting at that age. Although this is a non-negligible proportion, it is much smaller than the proportion that is removed annually at older ages, i.e., roughly $20-30 \%$ from age 3 years and onwards (Table 6.4.5.1.2). Although the Working Group did not run any simulations related to this problem, the catch of juvenile blue whiting is not likely to have a large effect on the dynamics of the stock. The catch of juvenile blue whiting in this fishery could probably be avoided only by strong reduction in effort, amounting to a practical closure of this component of the fishery.

The catch of 0 -group blue whiting by the Norwegian fleet was in the range of $4 \cdot 106$ to $120 \cdot 106$ individuals, which is several orders of magnitude less than the estimated stock numbers at the corresponding age. A catch of this magnitude is unlikely to have any noticeable effects on the stock as whole.

### 6.9 Quality of catch data and biological data

Without an internationally agreed TAC for this stock there are no obvious motives for mis-reporting. The extent of discarding in the fishery is not known. It is assumed that any discarding that occurs in the industrial fishery is confined to net bursting through overloading or through the catch exceeding the RSW tank capacity. On freezer vessels landing a catch for human consumption the role of grading in handling the catch is not known. If grading occurs in this sector then landing figures may not reflect the true catch. It would be helpful to the Working Group's deliberations if countries participating in the fishery could enumerate the number and category of vessel prosecuting the fishery.

The type and frequency of bycatch in the fishery has not been sufficiently explored. Each fishery could be expected to have a different bycatch profile. The submission of by catch data for each of the blue whiting fisheries would enable the WG to report on the likely impacts of blue whiting fisheries on other stocks and fisheries.

The text table below shows the number of samples and total landings by the three fisheries, Directed, Mixed, and Southern, and by quarter.

| Quarter |  | Directed | Mixed | Southern | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Number of samples | 99 | 8 | 127 | 234 |
|  | Landings (t) | 378827 | 13887 | 4987 | 397701 |
| 2 | Number of samples | 128 | 15 | 138 | 281 |
|  | Landings ( t ) | 504093 | 25451 | 5399 | 534943 |
| 3 | Number of samples | 159 | 42 | 116 | 317 |
|  | Landings (t) | 409960 | 54831 | 5111 | 469902 |
| 4 | Number of samples | 64 | 11 | 131 | 206 |
|  | Landings ( t ) | 106778 | 42002 | 3668 | 152448 |
| Total | Number of samples | 450 | 76 | 512 | 1038 |
|  | Landings (t) | 1399659 | 136171 | 19165 | 1554995 |

In total 1038 samples were collected from the fisheries, 117,090 fish measured and 19,551 fish aged. Sampled fish were not evenly distributed throughout the fisheries. The most intensive sampling took place in the south. Here one sample was taken for every 37 tonnes, followed by the mixed fishery with one sample for every 1792 tonnes, and lastly the directed fishery where there was one sample for every 3110 tonnes caught. In this context it should be noted that implementation of the EU Collection of Fisheries Data, Fisheries Regulation 1639/2001, requires a minimum of one sample to be taken for every 1000 t landed. Detailed information on the number of samples, number of fish measured, and number of fish aged by country and quarter is given in Table 6.9.1. As can be seen, no sampling was carried out by France, Germany, and Sweden and only limited sampling by Denmark, Ireland and Scotland.

The WG requires the samples to estimate catch in numbers and mean length and mean weight. Therefore, the WG urges all countries that exploit this stock to develop appropriate sampling schemes.

### 6.10 Recommendations

Several surveys on blue whiting are presently going on. It is recommended that a coordinated survey be organised covering the main spawning grounds of blue whiting. Other countries than those presently taking part in these surveys should be invited to take part.

It is recommended that information from existing surveys in which blue whiting are caught is made available to the Working Group. In particular, information from PGSPFN-coordinated surveys should be made available and analysed for information on abundance of incoming year classes.

It is suggested that the coordination could be taken care of by an extended ICES PGSPFN.

It would be helpful to the Working Group's deliberations if countries participating in the fishery could enumerate the number, capacity and effort of vessels prosecuting the fishery.

It is recommended that existing information on discards and bycatch in the fisheries is made available to the WG.

Table 6.2.1 Landings (tonnes) of BLUE WHITING from the directed fisheries (Sub-areas I and II, Division Va, XIVa and XIVb)
1987-2002, as estimated by the Working Group.

| Country | 1987 | 1988 | $1989{ }^{3)}$ | 1990 | 1991 | 1992 | 1993 | $1994{ }^{2)}$ | $1995{ }^{3)}$ | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |  |  |  |  |  |  |  |  | 15 | 7,721 | 5,723 | 13,608 |
| Estonia | - | - | - | - | - | - | - | - | - | 377 | 161 | 904 | - | - | - | - |
| Faroes | 9,290 | - | 1,047 | - | - | - | - | - | - | 345 | - | 44,594 | 11,507 | 17,980 | 64,496 | 82,977 |
| Germany | 1,010 | 3 | 1,341 | - | - | - | - | 2 | 3 | 32 | - | 78 | - | - | 3117 | 1,072 |
| Greenland | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |  |
| Iceland | - | - | 4,977 | - | - | - | - | - | 369 | 302 | 10,464 | 64,863 ${ }^{4}$ | 99,092 | 146,903 | 245,814 | 193,686 |
| Latvia | - | - | - | - | - | - | - | 422 | - | - | - | - | - | - | - | - |
| Netherlands | - | - | - | - | - | - | - | - | 72 | 25 | - | 63 | 435 | - | 5180 | 906 |
| Norway ${ }^{5}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 64,581 | 100,922 |
| Norway ${ }^{\text {a }}$ | - | - | - | 566 | 100 | 912 | 240 | - | - | 58 | 1,386 | 12,132 | 5,455 | - | 28,812 | - |
| Poland | 56 | 10 | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Sweden | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 850 |
| USSR/Russia ${ }^{\text {1) }}$ | 112,686 | 55,816 | 35,250 | 1,540 | 78,603 | 61,400 | 43,000 | 22,250 | 23,289 | 22,308 | 50,559 | 51,042 | 65,932 | 103,941 | 173,860 | 145,649 |
| Total | 123,042 | 55,829 | 42,615 | 2,106 | 78,703 | 62,312 | 43,240 | 22,674 | 23,733 | 23,447 | 62,570 | 173,676 | 182,436 | 276,545 | 591,583 | 539,670 |

${ }^{1}$ ) From 1992 only Russia
${ }^{2}$ ) Includes Vb for Russia.
${ }^{3}$ ) Icelandic mixed fishery in Va .
${ }^{4}$ ) include mixed in Va and directed in Vb .
${ }^{5)}$ Directed fishery
${ }^{6}$ By-catches of blue whiting in other fisheries.

Table 6.2.2 Landings (tonnes) of BLUE WHITING from directed fisheries (Division Vb,VIa,b, VIIb,c. VIIg-k and Sub-area XII) 1987-2002, as estimated by the Working Group.

| Country | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | $1998{ }^{\text {1) }}$ | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 2,655 | 797 | 25 | - | - | 3,167 | - | 770 | - | 269 | - | 5051 | 19,625 | 11,856 | 18,110 | 2,141 |
| Estonia | - | - | - | - | - | 6,156 | 1,033 | 4,342 | 7754 | 10,605 | 5,517 | 5,416 | - | - | - | - |
| Faroes | 70,625 | 79,339 | 70,711 | 43,405 | 10,208 | 12,731 | 14,984 | 22,548 | 26,009 | 18,258 | 22,480 | 26,328 | 93,234 | 129,969 | 188,464 | 115,127 |
| France | - | - | 2,190 | - | - | - | 1,195 | - | 720 | 6,442 | 12,446 | 7,984 | 6,662 | 13,481 | 13,480 | 14,688 |
| Germany | 3,850 | 5,263 | 4,073 | 1,699 | 349 | 1,307 | 91 | - | 6,310 | 6,844 | 4,724 | 17,891 | 3,170 | 12,655 | 15,862 | 15,378 |
| Iceland | - | - | - | - | - | - | - | - | - | - | - | - | 61,438 | 113,280 | 119,287 | 91,853 |
| Ireland | 3,706 | 4,646 | 2,014 | - | - | 781 | - | 3 | 222 | 1,709 | 25,785 | 45635 | 35,240 | 25,200 | 29,854 | 17,723 |
| Japan | - | - | - | - | - | 918 | 1,742 | 2,574 | - | - | - | - | - | - | - | - |
| Latvia | - | - | - | - | - | 10,742 | 10,626 | 2,160 | - | - | - | - | - | - | - | - |
| Lithauen | - | - | - | - | - | - | 2,046 | - | - | - | - | - | - | - | - | - |
| Netherlands ${ }^{2}$ ) | 5,627 | 800 | 2,078 | 7,280 | 17,359 | 11,034 | 18,436 | 21,076 | 26,703 | 17,644 | 23,676 | 27,884 | 35,408 | 46,128 | 68,415 | 33,365 |
| Norway | 191,012 | 208,416 | 258,386 | 281,036 | 114,866 | 148,733 | 198,916 | 226,235 | 261,272 | 337,434 | 318,531 | 519,622 | 475,004 | 460,274 | 399,932 | 385,495 |
| UK (Scotland) | 3,315 | 5,071 | 8,020 | 6,006 | 3,541 | 6,849 | 2,032 | 4,465 | 10,583 | 14,325 | 33,398 | 92,383 | 98,853 | 42,478 | 50,147 | 26,403 |
| Sweden | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 10 |
| $\underline{\text { USSR/Russia }{ }^{3} \text { ) }}$ | 165,497 | 121,705 | 127,682 | 124,069 | 72,623 | 115,600 | 96,000 | 94,531 | 83,931 | 64,547 | 68,097 | 79,000 | 112,247 | 141,257 | 141,549 | 144,419 |
| Total | 446,287 | 426,037 | 475,179 | 463,495 | 218,946 | 318,018 | 347,101 | 378,704 | 423,504 | 478,077 | 514,654 | 827,194 | 940,881 | 996,578 | 1,045,100 | 846,602 |

${ }^{1}$ ) Including some directed fishery also in Division IVa.
${ }^{2}$ ) Revised for the years 1987, 1988, 1989, 1992, 1995,1996, 1997
${ }^{3}$ ) From 1992 only Russia

Table 6.2.3 Landings (tonnes) of BLUE WHITING from directed fisheries and by-catches caught in other fisheries in Divisions IIIa, IVa 1987-2002, as estimated by the WG.

| Country | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | $1993{ }^{3)}$ | 1994 | 1995 | 1996 | 1997 | $1998{ }^{\text {2) }}$ | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark ${ }^{4)}$ |  |  | 3,632 | 10,972 | 5,961 | 4,438 | 25,003 | 5,108 | 4,848 | 29,137 | 9,552 | 40,143 | 36,492 | 30,360 | 21,995 | 35,530 |
| Denmark ${ }^{5}$ | 28,541 | 18,144 | 22,973 | 16,080 | 9,577 | 26,751 | 16,050 | 14,578 | 7,591 | 22,695 | 16,718 | 16,329 | 8,521 | 7,749 | 7,505 |  |
| Faroes ${ }^{4) 6}$ |  |  |  |  |  |  |  |  |  |  |  | - | - | - | 60 |  |
| $\text { Faroes }{ }^{5)^{6)}}$ | 7,051 | 492 | 3,325 | 5,281 | 355 | 705 | 1,522 | 1,794 | - | 6,068 | 6,066 | 296 | 265 | 42 | 6,741 | 7,317 |
| Germany ${ }^{1)}$ | 115 | 280 | 3 | - | - | 25 | 9 | - | - | - | - |  |  | - | 81 |  |
| Ireland | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 4 |
| Netherlands | - | - | - | 20 | - | 2 | 46 | - | - | - | 793 |  |  | - | - | 50 |
| $\text { Norway }^{4)}$ $\text { Norway }{ }^{5)}$ | 24,969 | 24,898 | 42,956 | 29,336 | 22,644 | 31,977 | 12,333 | 3,408 | 78,565 | 57,458 | 27,394 | 28,814 | 48,338 | 73,006 | $\begin{aligned} & 21,804 \\ & 58,182 \end{aligned}$ | 85,062 |
| Russia |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 69 |  |
| Sweden | 2,013 | 1,229 | 3,062 | 1,503 | 1,000 | 2,058 | 2,867 | 3,675 | 13,000 | 4,000 | 4,568 | 9,299 | 12,993 | 3,319 | 2,086 | 17,689 |
| UK | - | 100 | 7 | - | 335 | 18 | 252 | - | - | 1 | - |  |  | - | - |  |
| Total | 62,689 | 45,143 | 75,958 | 63,192 | 39,872 | 65,974 | 58,082 | 28,563 | 104,004 | 119,359 | 65,091 | 94,881 | 106,609 | 114,476 | 118,523 | 145,652 |
| ${ }^{1}$ ) Including directed fishery also in Division IVa. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{2}$ ) Including mixed industrial fishery in the Norwegian Sea |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{3}$ ) Imprecise estimates for Sweden: reported catch of 34265 t in 1993 is replaced by the mean of 1992 and 1994, i.e. 2,867 t , and used in the assessment. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{4)}$ Directed fishery |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{5}$ ) By-catches of blue whiting in other fisheries. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{6)}$ For the periode 1987-2000 landings figures also include landings from mixed fisheries in Division Vb . |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 6.2.4 Landings (tonnes) of BLUE WHITING from the Southern areas (Sub-areas VIII and IX and Divisions VIIg-k and VIId,e) 1987-2002, as estimated by the Working Group.

| Country | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Germany | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 600 |
| Ireland |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 98 |
| Netherlands | - | - | - | 450 | 10 | - | - | - | - | - | - | $10^{1)}$ | - | - | - | 3,208 |
| Norway | 4 | - | - | - | - | - | - | - | - | - | - |  |  | - | - |  |
| Portugal | 9,148 | 5,979 | 3,557 | 2,864 | 2,813 | 4,928 | 1,236 | 1,350 | 2,285 | 3,561 | 2,439 | 1,900 | 2,625 | 2,032 | 1,746 | 1,659 |
| Spain | 23,644 | 24,847 | 30,108 | 29,490 | 29,180 | 23,794 | 31,020 | 28,118 | 25,379 | 21,538 | 27,683 | 27,490 | 23,777 | 22,622 | 23,218 | 17,506 |
| UK | 23 | 12 | 29 | 13 | - | - | - | 5 | - | - | - | - | - | - | - |  |
| France | - | - | 1 | - | - | - | - | - | - | - | - | - | - | - | - |  |
| Total | 32,819 | 30,838 | 33,695 | 32,817 | 32,003 | 28,722 | 32,256 | 29,473 | 27,664 | 25,099 | 30,122 | 29,390 | 26,402 | 24,654 | 24,964 | 23,071 |

Table 6.2.5 Total landings of blue whiting by country and area for 2002 in tonnes. Landing figures provided by Working Group members and these figures may not be official catch statistics and therefore can not be used for management purposes.

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| I |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| IIa | 13,608 | 36,126 |  | 1,072 | 53,271 |  | 100,922 |  | 145,463 |  |  | 850 | 906 | 352,218 |
| IIIa | 6,454 |  |  |  |  |  |  |  |  |  |  | 17,610 |  | 24,064 |
| IVa | 28,621 | 7,163 |  |  |  | 4 | 85,062 |  |  |  |  | 79 | 50 | 120,979 |
| IVb | 455 | 154 |  |  |  |  |  |  |  |  |  |  |  | 609 |
| IXa |  |  |  |  |  |  |  | 1,659 |  |  |  |  |  | 1,659 |
| V |  |  |  |  |  |  |  |  | 107,900 |  |  |  |  | 107,900 |
| Va |  | 46,851 |  |  | 140,415 |  |  |  |  |  |  |  |  | 187,266 |
| Vb |  | 90,682 |  |  | 87,316 |  | 16,318 |  |  |  |  | 10 |  | 194,326 |
| VbVIVII |  |  | 14,688 |  |  |  |  |  |  |  |  |  |  | 14,688 |
| VIa | 1,428 | 1,315 |  | 8,598 |  | 11,394 | 105,434 |  |  | 4,135 |  |  | 12,099 | 144,403 |
| VIab+VIIbc |  |  |  |  | 1,915 |  |  |  |  |  |  |  |  | 1,915 |
| VIb |  | 22,739 |  | 500 |  |  | 203,133 |  |  |  |  |  | 4,104 | 230,476 |
| VIIb | 713 |  |  |  |  | 19 |  |  |  | 7,944 |  |  | 54 | 8,730 |
| VIIbe |  |  |  |  |  |  |  |  | 33,674 |  |  |  |  | 33,674 |
| VIIc |  |  |  | 6,280 |  | 6,310 | 41,121 |  |  | 14,324 |  |  | 17,108 | 85,143 |
| VIIgk+XII |  |  |  |  | 2,622 |  |  |  |  |  |  |  |  | 2,622 |
| VIIIabd |  |  |  |  |  |  |  |  |  |  |  |  | 3,203 | 3,203 |
| VIIIc+IXa |  |  |  |  |  |  |  |  |  |  | 17,506 |  |  | 17,506 |
| VIIIe |  |  |  |  |  | 35 |  |  |  |  |  |  |  | 35 |
| VIIj |  |  |  | 600 |  | 63 |  |  |  |  |  |  | 5 | 668 |
| VIIk |  |  |  |  |  |  | 13,509 |  |  |  |  |  |  | 13,509 |
| XII |  | 391 |  |  |  |  | 5,980 |  | 2,845 |  |  |  |  | 9,216 |
| Grand Total | 51,279 | 205,421 | 14,688 | 17,050 | 285,539 | 17,825 | 571,479 | 1,659 | 290,068 | 26,403 | 17,506 | 18,549 | 37,529 | 1,554,995 |

Table 6.2.6 Landings (tonnes) of BLUE WHITING from the main fisheries, 1987-2002, as estimated by the Working Group.

| 1987 | 123,042 |
| :--- | ---: |
| 1988 | 55,829 |
| 1989 | 42,615 |
| 1990 | 2,106 |
| 1991 | 78,703 |
| 1992 | 62,312 |
| 1993 | 43,240 |
| 1994 | 22,674 |
| 1995 | 23,733 |
| 1996 | 23,447 |
| 1997 | 62,570 |
| 1998 | 173,676 |
| 1999 | 182,436 |
| 2000 | 276,545 |
| 2001 | 591,583 |
| 2002 | 539,670 |

## Fishery in the spawning

 area (Divisions Vb, VIa, VIb and VIIb-c)Directed- and mixed fisheries (Divisions IIIa and IV )

Total northern areas
Total southern areas
(Subareas VIII and IX a (Subareas VIII and IX and
Divisions VIId, e, g-k)

| $\mathbf{6 3 2 , 0 1 8}$ | 32,819 |
| ---: | ---: |
| $\mathbf{5 2 7 , 0 0 9}$ | 30,838 |
| $\mathbf{5 9 3 , 7 5 2}$ | 33,695 |
| $\mathbf{5 2 8 , 7 9 3}$ | 32,817 |
| $\mathbf{3 3 7 , 5 2 1}$ | 32,003 |
| $\mathbf{4 4 6 , 3 6 7}$ | 28,722 |
| $\mathbf{4 4 8 , 4 2 3}$ | 32,256 |
| $\mathbf{4 2 9 , 9 4 1}$ | 29,473 |
| $\mathbf{5 5 1 , 2 4 1}$ | 27,664 |
| $\mathbf{6 2 0 , 8 8 3}$ | 25,099 |
| $\mathbf{6 4 2 , 3 1 5}$ | 30,122 |
| $\mathbf{1 , 0 9 5 , 7 5 1}$ | 29,400 |
| $\mathbf{1 , 2 2 9 , 9 2 6}$ | 26,402 |
| $\mathbf{1 , 3 8 7 , 5 9 9}$ | 24,654 |
| $\mathbf{1 , 7 5 5 , 2 0 6}$ | 24,964 |
| $\mathbf{1 , 5 1 5 , 7 9 3}$ | 39,202 |

Grand total

664,837
557,847
627,447
561,610
369,524
475,089
480,679
459,414
578,905
645,982
672,437
1,125,151
1,256,328
1,412,253
1,780,170
1,554,995

Table 6.2.7 Total landings of blue whiting by quarter and area for 2002 in tonnes. Landing figures provided by Working Group members.

| Area | Quarter 1 | Quarter 2 | Quarter 3 | Quarter 4 | Grand Total |
| :--- | ---: | ---: | ---: | ---: | ---: |
| I | 185 | 1 |  | 0 | 186 |
| IIa | 5,255 | 82,296 | 219,109 | 45,558 | 352,218 |
| IIIa | 1,796 | 3,387 | 12,140 | 6,741 | 24,064 |
| IVa | 19,541 | 27,045 | 47,910 | 26,483 | 120,979 |
| IVb |  | 3 | 493 | 113 | 609 |
| V | 15,832 | 45,575 | 2,857 | 43,636 | 107,900 |
| Va | 640 | 55,290 | 124,938 | 6,398 | 187,266 |
| Vb | 5,719 | 120,900 | 48,369 | 19,338 | 194,326 |
| VbVIVII | 4,565 | 5,307 | 4,320 | 496 | 14,688 |
| VIa | 2,526 | 141,853 | 7 | 17 | 144,403 |
| VIab+VIIbc | 1,335 | 580 |  |  | 1,915 |
| VIb | 210,817 | 19,659 |  |  | 230,476 |
| VIIb | 8,727 | 3 |  |  | 8,730 |
| VIIbc | 13,327 | 18,907 | 1,440 |  | 33,674 |
| VIIc | 76,479 | 8,664 |  |  | 85,143 |
| VIIgk+XII | 8,602 |  |  |  | 8,602 |
| VIIj | 645 | 18 |  |  | 668 |
| VIIk | 13,509 |  |  |  |  |
| VIIIabd |  |  |  | 3,203 |  |
| VIIIc+IXa | 4,713 | 4,827 | 4,525 | 3,442 | 17,506 |
| VIIIe | 35 |  |  |  | 35 |
| IXa | 274 | 572 | 586 | 226 | 1,659 |
| XII | 3,179 | 57 |  |  | 3,236 |
| Grand Total | 397,702 | 534,944 | 469,901 | 152,448 | $1,554,995$ |

Table 6.3.1.1 Blue whiting.Landings in numbers ('000) by length group (cm)
and quarters, quarters, for for the Northern area 2002

| length | Q1 | Q2 | Q3 | Q4 | All year |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0 | 0 | 0 |
| 8 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0 | 0 | 0 |
| 12 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 1 | 0 | 1 |
| 14 | 0 | 0 | 1 | 2 | 3 |
| 15 | 1 | 0 | 1 | 2 | 5 |
| 16 | 8 | 2 | 2 | 7 | 19 |
| 17 | 49 | 13 | 2 | 2 | 66 |
| 18 | 47 | 22 | 6 | 2 | 78 |
| 19 | 32 | 46 | 23 | 2 | 104 |
| 20 | 29 | 84 | 84 | 10 | 207 |
| 21 | 29 | 129 | 215 | 18 | 391 |
| 22 | 59 | 189 | 353 | 35 | 635 |
| 23 | 70 | 294 | 430 | 79 | 873 |
| 24 | 90 | 350 | 424 | 107 | 972 |
| 25 | 144 | 389 | 412 | 135 | 1079 |
| 26 | 179 | 357 | 393 | 139 | 1069 |
| 27 | 201 | 376 | 342 | 133 | 1053 |
| 28 | 240 | 365 | 244 | 93 | 943 |
| 29 | 261 | 309 | 194 | 66 | 830 |
| 30 | 257 | 282 | 133 | 35 | 707 |
| 31 | 199 | 184 | 93 | 10 | 487 |
| 32 | 144 | 121 | 50 | 13 | 327 |
| 33 | 113 | 81 | 15 | 5 | 213 |
| 34 | 75 | 55 | 11 | 4 | 145 |
| 35 | 61 | 33 | 5 | 3 | 101 |
| 36 | 25 | 23 | 4 | 0 | 53 |
| 37 | 18 | 15 | 1 | 5 | 39 |
| 38 | 12 | 13 | 1 | 2 | 28 |
| 39 | 9 | 7 | 0 | 2 | 18 |
| 40 | 2 | 2 | 1 | 5 | 10 |
| 41 | 1 | 2 | 0 | 0 | 2 |
| 42 | 1 | 1 | 0 | 0 | 2 |
| 43 | 0 | 0 | 0 | 0 | 0 |
| 44 | 0 | 0 | 0 | 0 | 0 |
| 45 | 0 | 0 | 0 | 0 | 0 |
| 46 | 0 | 0 | 0 | 0 | 0 |
| 47 | 0 | 0 | 0 | 0 | 0 |
| 48 | 0 | 0 | 0 | 0 | 0 |
| 49 | 0 | 0 | 0 | 0 | 0 |
| 50 | 0 | 0 | 0 | 0 | 0 |
| TOTAL numbers | 2355 | 3744 | 3444 | 919 | 10462 |
|  |  |  |  |  |  |
| Official Catch (t) | 364256 | 490720 | 404454 | 106284 | 1365714 |

Table 6.3.1.2 Blue whiting. Landings in numbers ('000) by length group (cm) and quarters for the North Sea and Skagerrak in 2002

| length | Q1 | Q2 | Q3 | Q4 | All year |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0 | 0 | 0 |
| 8 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0 | 0 | 0 |
| 12 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 2 | 0 | 2 |
| 14 | 0 | 0 | 33 | 8 | 41 |
| 15 | 1 | 0 | 57 | 14 | 73 |
| 16 | 3 | 0 | 14 | 3 | 20 |
| 17 | 20 | 1 | 7 | 2 | 29 |
| 18 | 19 | 6 | 0 | 2 | 28 |
| 19 | 15 | 24 | 9 | 2 | 50 |
| 20 | 17 | 53 | 66 | 4 | 141 |
| 21 | 20 | 47 | 163 | 28 | 259 |
| 22 | 23 | 46 | 126 | 52 | 247 |
| 23 | 32 | 41 | 78 | 65 | 217 |
| 24 | 30 | 34 | 80 | 80 | 224 |
| 25 | 22 | 29 | 75 | 63 | 189 |
| 26 | 12 | 25 | 51 | 61 | 150 |
| 27 | 12 | 12 | 25 | 34 | 83 |
| 28 | 5 | 9 | 19 | 16 | 50 |
| 29 | 4 | 4 | 13 | 8 | 29 |
| 30 | 5 | 3 | 5 | 13 | 26 |
| 31 | 1 | 2 | 3 | 4 | 10 |
| 32 | 1 | 3 | 3 | 3 | 10 |
| 33 | 1 | 0 | 0 | 1 | 2 |
| 34 | 1 | 0 | 0 | 0 | 2 |
| 35 | 0 | 0 | 1 | 0 | 2 |
| 36 | 1 | 0 | 0 | 0 | 1 |
| 37 | 0 | 0 | 0 | 0 | 0 |
| 38 | 0 | 0 | 0 | 0 | 0 |
| 39 | 0 | 0 | 0 | 0 | 0 |
| 40 | 0 | 0 | 0 | 0 | 0 |
| 41 | 0 | 0 | 0 | 0 | 0 |
| 42 | 0 | 0 | 0 | 0 | 0 |
| 43 | 0 | 0 | 0 | 0 | 0 |
| 44 | 0 | 0 | 0 | 0 | 0 |
| 45 | 0 | 0 | 0 | 0 | 0 |
| 46 | 0 | 0 | 0 | 0 | 0 |
| 47 | 0 | 0 | 0 | 0 | 0 |
| 48 | 0 | 0 | 0 | 0 | 0 |
| 49 | 0 | 0 | 0 | 0 | 0 |
| 50 | 0 | 0 | 0 | 0 | 0 |
| TOTAL numbers | 246 | 340 | 832 | 465 | 1883 |
|  |  |  |  |  |  |
| Official Catch (t) | 17070 | 23305 | 54796 | 40552 | 135723 |

Table 6.3.1.3 Blue whiting. Landings in numbers ('000) by length group (cm) and quarters for the Southern area in 2002

| length | Q1 | Q2 | Q3 | Q4 | All year |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0 | 0 | 0 |
| 8 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0 | 0 | 0 |
| 12 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 0 | 0 | 0 |
| 15 | 0 | 0 | 0 | 1 | 1 |
| 16 | 1 | 0 | 0 | 5 | 6 |
| 17 | 5 | 0 | 0 | 6 | 11 |
| 18 | 11 | 2 | 2 | 4 | 18 |
| 19 | 22 | 11 | 9 | 9 | 50 |
| 20 | 18 | 23 | 18 | 7 | 66 |
| 21 | 14 | 25 | 20 | 10 | 68 |
| 22 | 10 | 17 | 15 | 8 | 50 |
| 23 | 7 | 7 | 8 | 5 | 27 |
| 24 | 3 | 4 | 3 | 3 | 13 |
| 25 | 1 | 2 | 2 | 2 | 7 |
| 26 | 1 | 0 | 1 | 1 | 3 |
| 27 | 1 | 0 | 1 | 1 | 2 |
| 28 | 0 | 0 | 0 | 0 | 1 |
| 29 | 0 | 0 | 0 | 0 | 1 |
| 30 | 0 | 0 | 0 | 0 | 0 |
| 31 | 0 | 0 | 0 | 0 | 0 |
| 32 | 0 | 0 | 0 | 0 | 0 |
| 33 | 0 | 0 | 0 | 0 | 0 |
| 34 | 0 | 0 | 0 | 0 | 0 |
| 35 | 0 | 0 | 0 | 0 | 0 |
| 36 | 0 | 0 | 0 | 0 | 0 |
| 37 | 0 | 0 | 0 | 0 | 0 |
| 38 | 0 | 0 | 0 | 0 | 0 |
| 39 | 0 | 0 | 0 | 0 | 0 |
| 40 | 0 | 0 | 0 | 0 | 0 |
| 41 | 0 | 0 | 0 | 0 | 0 |
| 42 | 0 | 0 | 0 | 0 | 0 |
| 43 | 0 | 0 | 0 | 0 | 0 |
| 44 | 0 | 0 | 0 | 0 | 0 |
| 45 | 0 | 0 | 0 | 0 | 0 |
| 46 | 0 | 0 | 0 | 0 | 0 |
| 47 | 0 | 0 | 0 | 0 | 0 |
| 48 | 0 | 0 | 0 | 0 | 0 |
| 49 | 0 | 0 | 0 | 0 | 0 |
| 50 | 0 | 0 | 0 | 0 | 0 |
| TOTAL numbers | 93 | 92 | 80 | 63 | 328 |
|  |  |  |  |  |  |
| Official Catch (t) | 4988 | 5399 | 5110 | 3668 | 19165 |

Table 6.3.2.1 BLUE WHITING. Catch in number (millions) by age group in the directed fisheries (Sub-areas I and II, Divisions Va, and XIVa+b, Vb, VIa+b, VIIbc and VIIg-k) in 1991-2002.

| Age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 64 | - | - | - | 1 | 4 | 167 | 15 | 61 | 41 | 119 | 16 |
| 1 | 33 | 82 | 37 | 44 | 99 | 497 | 1352 | 984 | 544 | 912 | 3459 | 1111 |
| 2 | 533 | 52 | 130 | 31 | 143 | 327 | 1079 | 3535 | 1180 | 752 | 3924 | 2439 |
| 3 | 384 | 1509 | 335 | 190 | 338 | 451 | 751 | 3211 | 5257 | 3119 | 2728 | 2939 |
| 4 | 244 | 510 | 1348 | 362 | 416 | 425 | 526 | 929 | 3235 | 4834 | 3644 | 2114 |
| 5 | 330 | 200 | 376 | 1242 | 566 | 248 | 268 | 346 | 362 | 1517 | 2474 | 1804 |
| 6 | 235 | 139 | 196 | 294 | 769 | 430 | 238 | 311 | 186 | 500 | 555 | 1602 |
| 7 | 150 | 92 | 108 | 201 | 246 | 619 | 270 | 298 | 143 | 210 | 160 | 336 |
| 8 | 40 | 87 | 60 | 103 | 154 | 214 | 391 | 257 | 146 | 144 | 91 | 165 |
| 9 | 4 | 85 | 38 | 88 | 58 | 88 | 101 | 209 | 66 | 57 | 69 | 100 |
| $10+$ | 14 | 15 | 14 | 32 | 40 | 70 | 164 | 85 | 138 | 139 | 55 | 142 |
| Total | 2,032 | 2,770 | 2,641 | 2,588 | 2,829 | 3,373 | 5,307 | 10,180 | 11,318 | 12,225 | 17,281 | 142 |
| Tonnes $297,649379,549389,010401,378447,015493,373545,0581,000,8701,123,3171,273,1231,636,6831,399,659$ |  |  |  |  |  |  |  |  |  |  |  |  |

Table 6.3.2.2 BLUE WHITING. Catch in number (million) by age group in the directed fishery and bycatches from mixed fisheries (Divisions IIIa and IV) for 1991-2002.

| Age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 25 | - | 132 | 95 | 3303 | 812 | 29 | 11 | 60 | 56 | 9 | 190 |
| 1 | 8 | 160 | 167 | 33 | 101 | 1334 | 621 | 576 | 188 | 822 | 770 | 621 |
| 2 | 398 | 64 | 39 | 21 | 88 | 71 | 269 | 524 | 286 | 317 | 416 | 685 |
| 3 | 42 | 167 | 91 | 18 | 29 | 58 | 50 | 259 | 434 | 253 | 174 | 274 |
| 4 | 11 | 75 | 97 | 37 | 11 | 71 | 14 | 47 | 168 | 143 | 149 | 105 |
| 5 | 11 | 25 | 15 | 6 | 6 | 39 | 14 | 6 | 16 | 22 | 109 | 17 |
| 6 | 11 | 17 | 7 | 3 | 11 | 45 | 5 | 4 | 5 | 3 | 29 | 45 |
| 7 | 6 | 7 | 8 | 1 | 2 | 33 | 4 | 3 | 5 | 0 | 9 | 8 |
| 8 | 3 | 3 | - | 1 | 2 | 14 | 6 | 4 | 6 | 7 | 6 | 3 |
| 9 | 1 | 1 | - | 0 | 1 | 9 | 1 | 4 | 1 | 1 | 8 | 2 |
| $10+$ | 0 | 1 | - | - | 1 | 11 | 2 | 12 | 3 | 1 | 11 | 1 |
| Total | 518 | 519 | 556 | 214 | 3,555 | 2,499 | 1,015 | 1,450 | 1,172 | 1,627 | 1,689 | 1,951 |
| Tonnes 39,872 | 66,174 | 55,215 | 28,563 | $104,004119,35965,091$ | 94,881 | 106,609 | 114,477 | 118,523 | 136,171 |  |  |  |

Table 6.3.2.3 BLUE WHITING. Catch in number (millions) by age group in the Southern area, 1990-2001.

| Age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 70 | 19 | 25 | 13 | 3 | 9 | 11 | 18 | 18 | 32 | 33 | 17 |
| 1 | 181 | 139 | 41 | 12 | 96 | 43 | 118 | 97 | 57 | 80 | 134 | 88 |
| 2 | 182 | 205 | 146 | 56 | 123 | 131 | 143 | 122 | 82 | 123 | 146 | 108 |
| 3 | 70 | 95 | 181 | 149 | 55 | 117 | 86 | 71 | 130 | 93 | 60 | 79 |
| 4 | 39 | 43 | 62 | 72 | 38 | 36 | 26 | 69 | 57 | 35 | 14 | 24 |
| 5 | 17 | 12 | 12 | 27 | 44 | 33 | 8 | 32 | 35 | 9 | 10 | 4 |
| 6 | 8 | 6 | 7 | 9 | 20 | 17 | 4 | 7 | 15 | 10 | 1 | 1 |
| 7 | 3 | 2 | 2 | 5 | 6 | 5 | 3 | 2 | 3 | 3 | 0 | 0 |
| $8+$ | 3 | 1 | 1 | 4 | 5 | 3 | 3 | 4 | 2 | 0 | 0 | 0 |
| Total | 573 | 522 | 477 | 347 | 390 | 394 | 402 | 422 | 399 | 384 | 398 | 321 |
| Tonnes 32,003 | 28,722 | 32,256 | 29,468 | 27,664 | 25,099 | 30,122 | 29,400 | 26,402 | 24,654 | 24,964 | 19165 |  |


| Age | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 48 | 3512 | 437 | 584 | 1174 | 84 | 341 | 46 | 1949 | 83 | 161 |
| 1 | 258 | 148 | 2283 | 2291 | 1305 | 650 | 838 | 425 | 865 | 1611 | 267 |
| 2 | 348 | 274 | 567 | 2331 | 2044 | 816 | 578 | 721 | 718 | 703 | 1024 |
| 3 | 681 | 326 | 270 | 455 | 1933 | 1862 | 728 | 614 | 1340 | 672 | 514 |
| 4 | 334 | 548 | 286 | 260 | 303 | 1717 | 1897 | 683 | 791 | 753 | 302 |
| 5 | 548 | 264 | 299 | 285 | 188 | 393 | 726 | 1303 | 837 | 520 | 363 |
| 6 | 559 | 276 | 304 | 445 | 321 | 187 | 137 | 618 | 708 | 577 | 258 |
| 7 | 466 | 266 | 287 | 262 | 257 | 201 | 105 | 84 | 139 | 299 | 159 |
| 8 | 634 | 272 | 286 | 193 | 174 | 198 | 123 | 53 | 50 | 78 | 49 |
| 9 | 578 | 284 | 225 | 154 | 93 | 174 | 103 | 33 | 25 | 27 | 5 |
| $10+$ | 1460 | 673 | 334 | 255 | 259 | 398 | 195 | 50 | 38 | 95 | 10 |
| Total | 5914 | 6843 | 5578 | 7515 | 8051 | 6680 | 5771 | 4630 | 7460 | 5418 | 3112 |
| Tonnes | 909,556 | 576,419 | 570,072 | 641,766 | 695,596 | 826,986 | 664,837 | 557,847 | 627,447 | 561,610 | 369524 |


|  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| 0 | 19 | 198 | 42 | 3307 | 833 | 212 | 43 | 139 | 129 | 162 | 223 |
| 1 | 408 | 263 | 307 | 296 | 1893 | 2131 | 1657 | 788 | 1815 | 4364 | 1821 |
| 2 | 654 | 305 | 108 | 354 | 534 | 1519 | 4181 | 1549 | 1193 | 4486 | 3232 |
|  | 1642 | 621 | 368 | 422 | 632 | 904 | 3541 | 5821 | 3466 | 2962 | 3292 |
| 4 | 569 | 1571 | 389 | 465 | 537 | 578 | 1045 | 3461 | 5015 | 3807 | 2243 |
| 5 | 217 | 411 | 1222 | 616 | 323 | 296 | 384 | 413 | 1550 | 2593 | 1824 |
| 6 | 154 | 191 | 281 | 800 | 497 | 252 | 323 | 207 | 514 | 586 | 1647 |
| 7 | 110 | 107 | 174 | 254 | 663 | 282 | 303 | 151 | 213 | 170 | 344 |
| 8 | 80 | 65 | 90 | 160 | 232 | 407 | 264 | 153 | 151 | 97 | 169 |
| 9 | 32 | 38 | 79 | 60 | 98 | 104 | 212 | 69 | 58 | 77 | 103 |
| $10+$ | 12 | 17 | 31 | 42 | 83 | 169 | 86 | 141 | 140 | 66 | 143 |
| Total | 3896 | 3788 | 3091 | 6775 | 6327 | 6854 | 12039 | 12891 | 14244 | 19369 | 15040 |
| Tonnes | 475,089 | 480,679 | 459,414 | 578,905 | 645,982 | 672,437 | $1,125,151$ | $1,256,328$ | $1,412,253$ | $1,780,170$ | $1,554,995$ |

Table 6.3.3.1. Mean weights-at-age of blue whiting in the total catch in 1981-2002

| Age |  | $\mathbf{1 9 8 1}$ | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{0}$ | 0.038 | 0.018 | 0.020 | 0.026 | 0.016 | 0.030 | 0.023 | 0.031 | 0.014 | 0.034 |
|  | $\mathbf{1}$ | 0.052 | 0.045 | 0.046 | 0.035 | 0.038 | 0.040 | 0.048 | 0.053 | 0.059 | 0.045 |
|  | $\mathbf{2}$ | 0.065 | 0.072 | 0.074 | 0.078 | 0.074 | 0.073 | 0.086 | 0.076 | 0.079 | 0.070 |
|  | $\mathbf{3}$ | 0.103 | 0.111 | 0.118 | 0.089 | 0.097 | 0.108 | 0.106 | 0.097 | 0.103 | 0.106 |
|  | $\mathbf{4}$ | 0.125 | 0.143 | 0.140 | 0.132 | 0.114 | 0.130 | 0.124 | 0.128 | 0.126 | 0.123 |
|  | $\mathbf{5}$ | 0.141 | 0.156 | 0.153 | 0.153 | 0.157 | 0.165 | 0.147 | 0.142 | 0.148 | 0.147 |
|  | $\mathbf{6}$ | 0.155 | 0.177 | 0.176 | 0.161 | 0.177 | 0.199 | 0.177 | 0.157 | 0.158 | 0.168 |
|  | $\mathbf{7}$ | 0.170 | 0.195 | 0.195 | 0.175 | 0.199 | 0.209 | 0.208 | 0.179 | 0.171 | 0.175 |
|  | $\mathbf{8}$ | 0.178 | 0.200 | 0.200 | 0.189 | 0.208 | 0.243 | 0.221 | 0.199 | 0.203 | 0.214 |
|  | $\mathbf{9}$ | 0.187 | 0.204 | 0.204 | 0.186 | 0.218 | 0.246 | 0.222 | 0.222 | 0.224 | 0.217 |
| $\mathbf{1 0 +}$ | 0.213 | 0.231 | 0.228 | 0.206 | 0.237 | 0.257 | 0.254 | 0.260 | 0.253 | 0.256 | 0.266 |
|  |  |  |  |  |  |  |  |  |  |  |  |
| $\mathbf{A g e}$ |  | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ |
|  | $\mathbf{0}$ | 0.024 | 0.028 | 0.033 | 0.022 | 0.018 | 0.031 | 0.033 | 0.035 | 0.031 | 0.038 |
|  | $\mathbf{1}$ | 0.057 | 0.066 | 0.061 | 0.064 | 0.041 | 0.047 | 0.048 | 0.063 | 0.057 | 0.050 |
|  | $\mathbf{2}$ | 0.083 | 0.082 | 0.087 | 0.091 | 0.080 | 0.072 | 0.072 | 0.078 | 0.075 | 0.078 |
|  | $\mathbf{3}$ | 0.119 | 0.109 | 0.108 | 0.118 | 0.102 | 0.102 | 0.094 | 0.088 | 0.086 | 0.094 |
|  | $\mathbf{4}$ | 0.140 | 0.137 | 0.137 | 0.143 | 0.116 | 0.121 | 0.125 | 0.109 | 0.104 | 0.108 |
|  | $\mathbf{5}$ | 0.167 | 0.163 | 0.164 | 0.154 | 0.147 | 0.140 | 0.149 | 0.142 | 0.133 | 0.129 |
|  | $\mathbf{6}$ | 0.193 | 0.177 | 0.189 | 0.167 | 0.170 | 0.166 | 0.178 | 0.170 | 0.156 | 0.163 |
|  | $\mathbf{7}$ | 0.226 | 0.200 | 0.207 | 0.203 | 0.214 | 0.177 | 0.183 | 0.199 | 0.179 | 0.186 |
|  | $\mathbf{8}$ | 0.235 | 0.217 | 0.217 | 0.206 | 0.230 | 0.183 | 0.188 | 0.193 | 0.187 | 0.193 |
| $\mathbf{9}$ | 0.284 | 0.225 | 0.247 | 0.236 | 0.238 | 0.203 | 0.221 | 0.192 | 0.232 | 0.231 | 0.233 |
| $\mathbf{1 0 +}$ | 0.294 | 0.281 | 0.254 | 0.256 | 0.279 | 0.232 | 0.248 | 0.245 | 0.241 | 0.243 | 0.262 |

Table 6.3.4.1. Proportion of mature blue whiting at age. The values were estimated by the 1994 WG (ICES 1995/Assess:7).

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Proportion <br> of mature | 0.00 | 0.11 | 0.40 | 0.82 | 0.86 | 0.91 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |

Table 6.4.1.1 Age-stratified acoustic survey estimates of blue whiting in the spawning area west of the British Isles
R.V."Hjort" March/ April 1981-2003. Number in millions

| Numbers | age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | total |
| 1981 |  | 2372 | 7583 | 3253 | 3647 | 4611 | 4638 | 3654 | 2591 | 1785 | 34134 |
| 1982 |  | no survey |  |  |  |  |  |  |  |  |  |
| 1983 |  | 297 | 2108 | 2723 | 6511 | 3735 | 3650 | 3153 | 2279 | 1182 | 25638 |
| 1984 |  | 15767 | 1721 | 1616 | 1719 | 1858 | 1128 | 567 | 440 | 348 | 25164 |
| 1985 |  | no survey |  |  |  |  |  |  |  |  |  |
| 1986 |  | 1003 | 5829 | 4122 | 624 | 228 | 203 | 250 | 137 | 170 | 12566 |
| 1987 |  | 4960 | 8417 | 22589 | 4735 | 282 | 417 | 385 | 159 | 27 | 41971 |
| 1988 |  | 9712 | 9090 | 12367 | 20392 | 7355 | 723 | 599 | 326 | 398 | 60962 |
| 1989 |  | 6787 | 22270 | 9973 | 10504 | 7803 | 933 | 293 | 177 | 46 | 58786 |
| 1990 |  | 14169 | 12670 | 11228 | 5587 | 6556 | 3273 | 516 | 183 | 108 | 54290 |
| 1991 |  | 11147 | 6340 | 8497 | 7407 | 4558 | 2019 | 545 | 96 | 16 | 40625 |
| 1992 |  | 1232 | 26123 | 4719 | 1574 | 1386 | 810 | 616 | 257 | 19 | 36736 |
| 1993 |  | 4489 | 3321 | 26771 | 2643 | 1270 | 557 | 426 | 108 | 22 | 39607 |
| 1994 |  | 1603 | 2950 | 4476 | 11354 | 1742 | 1687 | 908 | 770 | 207 | 25697 |
| 1995 |  | 8538 | 9874 | 7906 | 6861 | 9467 | 1795 | 1083 | 482 | 149 | 46155 |
| 1996 |  | 8781 | 7433 | 8371 | 2399 | 4455 | 4111 | 1202 | 459 | 162 | 37373 |
| 1997 |  | no survey |  |  |  |  |  |  |  |  |  |
| 1998 |  | 18218 | 34991 | 4697 | 1674 | 279 | 407 | 381 | 351 | 86 | 61084 |
| 1999 |  | 19034 | 60309 | 26103 | 1481 | 316 | 72 | 153 | 141 | 0 | 107609 |
| 2000 |  | 8613 | 31011 | 41382 | 6843 | 898 | 427 | 228 | 139 | 115 | 89656 |
| 2001 |  | 44162 | 12843 | 13805 | 8292 | 718 | 175 | 51 | 0 | 0 | 80046 |
| 2002 | 20455 | 71996 | 54740 | 12757 | 5266 | 8404 | 1450 | 305 | 15 | 176 | 175564 |
| 2003 |  | 23992 | 70303 | 28756 | 5735 | 2430 | 1708 | 260 |  |  | 133184 |

Table 6.4.1.2. Age-stratified acoustic survey estimates of blue whiting in the spawning area by Russian vessels
Number in millions,

| Numbers | age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | total |
| 1981 |  |  |  |  |  |  |  |  |  |  | 0 |
| 1982 |  |  | 540 | 2750 | 1340 | 1380 | 1570 | 2350 | 1730 | 1290 |  |
| 1983 |  |  | 2330 | 2930 | 9390 | 3880 | 1970 | 1370 | 780 | 660 | 23310 |
| 1984 |  |  | 2900 | 800 | 1100 | 4200 | 2200 | 1200 | 1700 | 1200 | 15300 |
| 1985 |  |  | 13220 | 930 | 580 | 1780 | 860 | 610 | 580 | 540 |  |
| 1986 |  |  | 18750 | 23180 | 2540 | 610 | 620 | 750 | 640 | 710 | 47800 |
| 1987 |  |  | 4480 | 19170 | 5860 | 1070 | 500 | 810 | 860 | 670 | 33420 |
| 1988 |  |  | 3710 | 4550 | 8610 | 4130 | 1270 | 480 | 250 | 260 | 23260 |
| 1989 |  |  | 11910 | 7120 | 6670 | 6970 | 4580 | 2750 | 1880 | 810 | 42690 |
| 1990 |  |  | 9740 | 12140 | 5740 | 2580 | 1470 | 220 | 80 | 10 | 31980 |
| 1991 |  |  | 10300 | 5350 | 5130 | 2630 | 1770 | 870 | 300 | 220 | 26570 |
| 1992 |  |  | 20010 | 6700 | 1350 | 440 | 390 | 170 | 0 | 0 | 29060 |
| 1993 |  |  | 4728 | 12337 | 5304 | 2249 | 1316 | 621 | 386 | 150 | 27091 |
| 1994 |  |  | no survey |  |  |  |  |  |  |  | 0 |
| 1995 |  |  | 12657 | 10028 | 8942 | 2651 | 1093 | 408 | 131 | 14 | 35924 |
| 1996 |  |  | 15285 | 10629 | 4897 | 6940 | 1482 | 653 | 85 | 0 | 39971 |
| 1997 |  |  | no survey |  |  |  |  |  |  |  |  |
| 1998 |  |  | no survey |  |  |  |  |  |  |  | 0 |
| 1999 |  |  | no survey |  |  |  |  |  |  |  | 0 |
| 2000 |  |  | no survey |  |  |  |  |  |  |  | 0 |
| 2001 |  |  | no survey |  |  |  |  |  |  |  | 0 |
| 2002 |  |  |  |  |  |  |  |  |  |  | 0 |
| 2003 |  |  |  |  |  |  |  |  |  |  | 0 |
|  |  |  |  | d in the | essmen |  |  |  |  |  |  |

Table 6.4.1.3. Age-stratified acoustic survey estimates of blue whiting in the Norwegian Sea in July
Number in millions

| Numbers | age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | total |
| 1981 |  | 182 | 728 | 4542 | 3874 | 2678 | 2834 | 2964 | 2756 | 2054 | 22612 |
| 1982 |  | 184 | 460 | 1242 | 4715 | 3611 | 3128 | 2323 | 1679 | 874 | 18216 |
| 1983 |  | 22356 | 396 | 468 | 756 | 1404 | 576 | 468 | 432 | 324 | 27180 |
| 1984 |  | 30380 | 13916 | 833 | 392 | 539 | 539 | 343 | 49 | 49 | 47040 |
| 1985 |  | 5969 | 23876 | 12502 | 658 | 423 | 188 | 235 | 141 | 376 | 44368 |
| 1986 |  | 2324 | 2380 | 7224 | 6944 | 1876 | 952 | 336 | 308 | 140 | 22484 |
| 1987 |  | 8204 | 4032 | 5180 | 5572 | 1204 | 224 | 168 | 56 | 84 | 24724 |
| 1988 |  | 4992 | 2880 | 2640 | 3480 | 912 | 120 | 96 | 24 | 48 | 15192 |
| 1989 |  | 1172 | 1125 | 812 | 379 | 410 | 212 | 22 | 32 |  | 4164 |
| 1990 |  | no survey |  |  |  |  |  |  |  |  |  |
| 1991 |  | no survey |  |  |  |  |  |  |  |  |  |
| 1992 |  | 792 | 1134 | 6939 | 766 | 247 | 172 | 90 | 11 | 18 | 10169 |
| 1993 |  | 830 | 125 | 1070 | 6392 | 1222 | 489 | 248 | 58 | 88 | 10522 |
| 1994 |  | no survey |  |  |  |  |  |  |  |  |  |
| 1995 |  | 6974 | 2811 | 1999 | 1209 | 1622 | 775 | 173 | 61 |  | 15624 |
| 1996 |  | 23464 | 1057 | 899 | 649 | 436 | 505 | 755 | 69 | 41 | 27875 |
| 1997 |  | 30227 | 25638 | 1524 | 779 | 300 | 407 | 260 | 137 | 123 | 59395 |
| 1998 |  | 24244 | 47815 | 16282 | 556 | 212 | 100 | 64 | 10 | 255 | 89538 |
| 1999 |  | 14367 | 9750 | 23701 | 9754 | 1733 | 466 | 79 | 48 | 91 | 59989 |
| 2000 |  | 25813 | 3298 | 2721 | 3078 | 23 | 46 | 6 |  |  | 34985 |
| 2001 |  | 61470 | 22051 | 7883 | 3225 | 1824 | 156 | 12 | 0 | 68 | 96689 |
| 2002 |  |  | survey |  |  |  |  |  |  |  | 0 |
| 2003 |  |  |  |  |  |  |  |  |  |  | 0 |

Table 6.4.1.1.1 Age- and length distribution of blue whiting in the survey by R.V. "Johan Hjort", west of the British Isles, April 2003.

| Length (cm) | Age in years |  |  |  |  |  |  |  |  | Numbers$\left(10^{6}\right)$ | Biomass$\left(10^{6} \mathrm{~kg}\right)$ | Mean weigth, (g) | Proportionmature |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{array}{r} 1 \\ 2002 \\ \hline \end{array}$ | $\begin{array}{r} 2 \\ 2001 \\ \hline \end{array}$ | $\begin{array}{r} 3 \\ 2000 \\ \hline \end{array}$ | $\begin{array}{r} 4 \\ 1999 \\ \hline \end{array}$ | 5 1998 | $\begin{array}{r} 6 \\ 1997 \\ \hline \end{array}$ | $\begin{array}{r} 7 \\ 1996 \\ \hline \end{array}$ | $\begin{array}{r} 8 \\ 1995 \\ \hline \end{array}$ | $\begin{array}{r} 9 \\ 1994 \\ \hline \end{array}$ |  |  |  |  |
| 14.0-15.0 | 167 |  |  |  |  |  |  |  |  | 167 | 2.8 | 16.8 | 2.9 |
| 15.0-16.0 | 1015 |  |  |  |  |  |  |  |  | 1015 | 19.4 | 19.1 | 4.1 |
| 16.0-17.0 | 3105 |  |  |  |  |  |  |  |  | 3105 | 68.1 | 21.9 | 5.7 |
| 17.0-18.0 | 3939 |  |  |  |  |  |  |  |  | 3939 | 108.1 | 27.4 | 7.9 |
| 18.0-19.0 | 7135 | 18 |  |  |  |  |  |  |  | 7153 | 236.5 | 33.1 | 11.0 |
| 19.0-20.0 | 6853 | 45 |  |  |  |  |  |  |  | 6898 | 262.7 | 38.1 | 15.0 |
| 20.0-21.0 | 3159 | 812 |  |  |  |  |  |  |  | 3971 | 174.4 | 43.9 | 31.5 |
| 21.0-22.0 | 1042 | 1983 | 325 |  |  |  |  |  |  | 3349 | 163.0 | 48.7 | 65.1 |
| 22.0-23.0 | 106 | 4995 | 2084 |  |  |  |  |  |  | 7185 | 399.5 | 55.6 | 86.9 |
| 23.0-24.0 |  | 6600 | 7823 | 45 |  |  |  |  |  | 14468 | 897.4 | 62.0 | 92.0 |
| 24.0-25.0 |  | 7175 | 17435 | 1395 |  |  |  |  |  | 26004 | 1758.6 | 67.6 | 95.4 |
| 25.0-26.0 |  | 1834 | 22266 | 5018 | 105 |  |  |  |  | 29223 | 2184.4 | 74.7 | 97.9 |
| 26.0-27.0 |  | 530 | 13601 | 6551 | 366 | 172 |  |  |  | 21220 | 1758.5 | 82.9 | 99.1 |
| 27.0-28.0 |  |  | 4548 | 7756 | 1699 | 101 | 203 |  |  | 14307 | 1314.0 | 91.8 | 99.8 |
| 28.0-29.0 |  |  | 1755 | 3557 | 1050 | 953 |  |  |  | 7316 | 751.6 | 102.7 | 99.9 |
| 29.0-30.0 |  |  | 466 | 3321 | 865 | 301 | 956 | 233 |  | 6143 | 694.2 | 113.0 | 100 |
| 30.0-31.0 |  |  |  | 802 | 1025 | 70 |  |  | 150 | 2048 | 266.5 | 130.1 | 100 |
| 31.0-32.0 |  |  |  | 311 | 605 | 160 |  |  |  | 1076 | 149.1 | 138.5 | 100 |
| 32.0-33.0 |  |  |  |  |  | 364 | 351 |  |  | 715 | 113.3 | 158.5 | 100 |
| 33.0-34.0 |  |  |  |  |  | 223 | 121 |  |  | 344 | 67.7 | 196.9 | 100 |
| 34.0-35.0 |  |  |  |  | 20 | 85 | 51 |  |  | 156 | 26.8 | 172.5 | 100 |
| 35.0-36.0 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 36.0-37.0 |  |  |  |  |  |  | 27 | 27 | 79 | 133 | 28.1 | 212.1 | 100 |
| TSN ( $10^{6}$ ) | 26520 | 23992 | 70303 | 28756 | 5735 | 2430 | 1708 | 260 | 229 | 159935 |  |  |  |
| TSB ( $10^{6} \mathrm{~kg}$ ) | 895 | 1487 | 5220 | 2637 | 616 | 303 | 218 | 32 | 37 | 11445 |  |  |  |
| Mean length (cm) | 18.6 | 23.5 | 25.3 | 27.3 | 28.9 | 30.0 | 30.4 | 30.2 | 32.6 | 24.6 |  |  |  |
| Mean weight (g) | 33.7 | 62.0 | 74.3 | 91.7 | 107.5 | 124.5 | 127.6 | 121 | 161.8 | 71.6 |  |  |  |
| Condition | 5.2 | 4.8 | 4.6 | 4.5 | 4.5 | 4.6 | 4.5 | 4.4 | 4.7 | 4.8 |  |  |  |
| \% mature | 10.1 | 77.4 | 99.1 | 100 | 100 | 100 | 100 | 100 | 100 | 92.0 |  |  |  |
| \% of SSB | 0.9 | 11.2 | 50.4 | 25.7 | 6.0 | 3.0 | 2.1 | 0.3 | 0.3 |  |  |  |  |

Table 6.4.1.1.2 Total stock estimate of blue whiting, spring 2003, as estimated from Russian surveys with R.V. "Atlantniro" and R.V. "Smolensk".

| Length, cm | AGE in years |  |  |  |  |  |  |  |  |  |  | Number$10^{6} \text { ind }$ | Biomass$10^{3} \mathrm{t}$ | mean weight, gr | CF |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |  |  |  |  |
|  | 2002 | 2001 | 2000 | 1999 | 1998 | 1997 | 1996 | 1995 | 1994 | 1993 | 1992 |  |  |  |  |
| 14.0-15.0 | 4.9 |  |  |  |  |  |  |  |  |  |  | 4.9 | 0.1 | 18.5 | 6.7 |
| 15.0-16.0 | 635.8 |  |  |  |  |  |  |  |  |  |  | 635.8 | 9.9 | 15.6 | 4.6 |
| 16.0-17.0 | 3291.3 |  |  |  |  |  |  |  |  |  |  | 3291.3 | 65.3 | 19.8 | 4.8 |
| 17.0-18.0 | 9779.4 | 18.2 |  |  |  |  |  |  |  |  |  | 9797.6 | 230.6 | 23.5 | 4.8 |
| 18.0-19.0 | 6294.3 | 31.6 |  |  |  |  |  |  |  |  |  | 6325.9 | 179.7 | 28.4 | 4.9 |
| 19.0-20.0 | 4602.2 | 488.9 |  |  |  |  |  |  |  |  |  | 5091.1 | 169.7 | 33.3 | 4.9 |
| 20.0-21.0 | 3208.7 | 823.1 |  |  |  |  |  |  |  |  |  | 4031.8 | 153.4 | 38.0 | 4.8 |
| 21.0-22.0 | 1030.3 | 881.6 | 200.6 |  |  |  |  |  |  |  |  | 2112.5 | 99.0 | 46.8 | 5.1 |
| 22.0-23.0 | 1121.6 | 2114.6 | 486.4 |  |  |  |  |  |  |  |  | 3722.6 | 196.9 | 52.9 | 5.0 |
| 23.0-24.0 | 522.1 | 5039.0 | 5597.3 | 379.0 |  |  |  |  |  |  |  | 11537.5 | 677.1 | 58.7 | 4.8 |
| 24.0-25.0 | 531.4 | 8874.3 | 15512.9 | 3909.3 | 220.4 |  |  |  |  |  |  | 29048.3 | 1902.7 | 65.5 | 4.7 |
| 25.0-26.0 |  | 5675.5 | 23941.7 | 10688.2 | 1877.4 | 169.2 |  |  |  |  |  | 42352.0 | 3143.6 | 74.2 | 4.8 |
| 26.0-27.0 |  | 1532.1 | 13539.9 | 17183.8 | 5664.2 | 604.6 | 101.2 |  |  |  |  | 38625.7 | 3171.5 | 82.1 | 4.7 |
| 27.0-28.0 |  | 83.8 | 4915.0 | 14153.4 | 7582.3 | 1442.8 | 283.5 |  |  |  |  | 28460.8 | 2605.1 | 91.5 | 4.7 |
| 28.0-29.0 |  | 230.2 | 722.9 | 6643.0 | 6792.1 | 2720.8 | 582.8 | 216.5 | 105.6 |  |  | 18013.8 | 1862.8 | 103.4 | 4.7 |
| 29.0-30.0 |  |  |  | 2176.2 | 4699.2 | 1946.0 | 1668.3 | 446.4 | 81.6 |  |  | 11017.7 | 1254.3 | 113.8 | 4.7 |
| 30.0-31.0 |  |  | 125.9 | 418.3 | 1820.1 | 2665.7 | 1576.7 | 299.7 | 70.7 |  |  | 6977.0 | 893.0 | 128.0 | 4.7 |
| 31.0-32.0 |  |  |  | 82.3 | 1277.2 | 1991.7 | 759.7 | 306.7 | 71.5 |  |  | 4489.0 | 646.0 | 143.9 | 4.8 |
| 32.0-33.0 |  |  |  | 25.8 | 543.5 | 927.7 | 984.1 | 416.9 |  |  |  | 2898.1 | 452.4 | 156.1 | 4.8 |
| 33.0-34.0 |  |  |  |  | 134.0 | 407.8 | 324.9 | 427.6 | 65.5 | 33.6 |  | 1393.3 | 255.2 | 183.1 | 5.1 |
| 34.0-35.0 |  |  |  |  | 12.5 | 143.8 | 173.0 | 207.4 | 126.0 | 39.5 |  | 702.2 | 140.8 | 200.6 | 5.1 |
| 35.0-36.0 |  |  |  |  | 7.6 | 116.0 | 146.3 | 94.9 | 131.6 | 67.6 |  | 564.1 | 123.1 | 218.2 | 5.1 |
| 36.0-37.0 |  |  |  |  | 9.4 | 7.4 | 91.4 | 136.6 | 145.6 | 15.0 | 65.3 | 470.7 | 115.5 | 245.4 | 5.3 |
| 37.0-38.0 |  |  |  |  |  | 24.7 | 16.5 | 103.7 | 50.8 | 136.4 | 51.3 | 383.5 | 95.3 | 248.5 | 4.9 |
| 38.0-39.0 |  |  |  |  |  |  | 10.7 | 51.6 | 43.7 | 42.5 | 11.9 | 160.4 | 49.1 | 306.2 | 5.6 |
| 39.0-40.0 |  |  |  |  |  |  | 6.6 | 19.8 | 9.1 | 5.1 |  | 40.6 | 12.5 | 307.7 | 5.2 |
| 40.0-41.0 |  |  |  |  |  |  | 4.2 | 4.2 | 11.4 |  | 11.5 | 31.3 | 10.4 | 331.7 | 5.2 |
| 41.0-42.0 |  |  |  |  |  |  |  |  |  | 5.0 | 9.2 | 14.1 | 4.9 | 345.2 | 5.0 |
| 42.0-43.0 |  |  |  |  |  |  |  |  |  | 52.1 | 3.0 | 55.1 | 22.2 | 402.6 | 5.4 |
| 43.0-44.0 |  |  |  |  |  |  |  |  |  |  | 4.3 | 4.3 | 1.9 | 440.0 | 5.5 |
| Number, $10{ }^{6}$ | 31022.0 | 25792.9 | 65042.6 | 55659.3 | 30639.8 | 13168.3 | 6729.8 | 2732.0 | 913.0 | 396.7 | 156.5 | 232252.8 |  |  |  |
| $<\mathrm{L}>$, cm | 18.2 | 23.7 | 25.0 | 26.3 | 27.7 | 29.4 | 30.3 | 31.8 | 33.2 | 36.8 | 37.4 | 25.14 |  |  |  |
| Biomass, $10^{3} \mathrm{t}$ | 953.6 | 1708.2 | 4760.5 | 4789.7 | 3077.8 | 1609.1 | 885.6 | 434.4 | 176.2 | 105.1 | 43.6 |  | 18544.0 |  |  |
| <weight>, gr | 30.7 | 66.2 | 73.2 | 86.1 | 100.5 | 122.2 | 131.6 | 159.0 | 193.0 | 264.9 | 278.9 |  |  | 79.8 |  |
| Spawning <br> Number, $10^{6}$ | 4837.0 | 15306.8 | 58702.7 | 54485.1 | 30471.9 | 13168.3 | 6729.8 | 2732.0 | 913.0 | 396.7 | 156.5 | 187899.8 |  |  |  |
| Biomass, $10^{3} \mathrm{t}$ | 223.6 | 1103.1 | 4294.7 | 4700.9 | 3055.7 | 1608.9 | 885.6 | 434.4 | 176.2 | 105.1 | 43.6 |  | 16631.9 | 88.5 |  |
| CF | 5.1 | 5.0 | 4.7 | 4.7 | 4.7 | 4.8 | 4.7 | 4.9 | 5.3 | 5.3 | 5.3 |  |  |  | 4.7 |

Table 6.4.1.1.3. Blue whiting biomass estimates (million tonnes) in the spawning area.

| Year | Russia total | Russia spawning | Norway total | Norway spawning | Faroes total | Faroes spawning |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1983 | 3.6 | 3.6 | 4.7 | 4.4 |  |  |
| 1984 | 3.4 | 2.7 | 2.8 | 2.1 | 2.4 | 2.2 |
| 1985 | 2.8 | 2.7 |  |  | 6.4 | 1.7 |
| 1986 | 6.4 | 5.6 | 2.6 | 2.0 |  |  |
| 1987 | 5.4 | 5.1 | 4.3 | 4.1 |  |  |
| 1988 | 3.7 | 3.1 | 7.1 | 6.8 |  |  |
| 1989 | 6.3 | 5.7 | 7.0 | 6.1 |  |  |
| 1990 | 5.4 | 5.1 | 6.3 | 5.7 |  |  |
| 1991 | 4.6 | 4.2 | 5.1 | 4.8 |  |  |
| 1992 | 3.6 | 3.3 | 4.3 | 4.2 |  |  |
| 1993 | 3.8 | 3.7 | 5.2 | 5.0 |  |  |
| 1994 |  |  | 4.1 | 4.1 |  |  |
| 1995 | 6.8 | 6.0 | 6.7 | 6.1 |  |  |
| 1996 | 7.1 | 5.8 | 5.1 | 4.5 |  |  |
| 1997 |  |  |  |  |  |  |
| 1998 |  |  | 5.5 | 4.7 |  |  |
| 1999 |  |  | 8.9 | 8.5 |  |  |
| 2000 |  |  | 8.3 | 7.8 |  |  |
| 2001 |  |  | 6.7 | 5.6 |  |  |
| 2002 | 5.2 |  | 12.2 | 10.9 |  |  |
| 2003 | 18.5 | 16.6 | 11.4 | 10.4 |  |  |
| Mean | 5.8 | 5.2 | 6.2 | 5.7 | 4.4 | 2.0 |

Table 6.4.1.2.1 Age stratified acoustic survey estimates of blue whiting in the Icelandic EEZ in July Numbers in millions, weight in thousand tonnes Mean length in cm .

| Numbers | Age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ | total |
| 1999 | 14869 | 2100 | 1357 | 1772 | 5790 | 1344 | 316 | 50 | 15 | 42 | 27655 |
| 2000 | 10683 | 8594 | 934 | 523 | 1218 | 468 | 106 | 25 | 1 | 1 | 22553 |
| 2001 | 27305 | 4090 | 5215 | 1657 | 1614 | 398 | 132 | 37 | 6 | 2 | 40456 |
| 2002 | 3815 | 10785 | 3107 | 1436 | 1724 | 1430 | 727 | 178 | 47 | 5 | 23254 |
| Biomass |  |  |  |  |  |  |  |  |  |  |  |
| 1999 | 265 | 163 | 127 | 201 | 764 | 212 | 55 | 13 | 4 | 14 | 1818 |
| 2000 | 186 | 624 | 85 | 63 | 167 | 78 | 22 | 5 |  |  | 1230 |
| 2001 | 661 | 295 | 568 | 211 | 231 | 66 | 22 | 8 | 1 |  | 2063 |
| 2002 | 77 | 746 | 297 | 160 | 217 | 203 | 114 | 31 | 13 | 1 | 1859 |
| length |  |  |  |  |  |  |  |  |  |  |  |
| 1999 | 13.5 | 23.5 | 25.0 | 26.3 | 27.6 | 29.0 | 30.3 | 33.7 | 35.5 | 36.8 | 19.7 |
| 2000 | 13.5 | 22.6 | 24.4 | 26.2 | 27.6 | 29.6 | 32.3 | 32.0 | 36.0 | 28.0 | 18.8 |
| 2001 | 15.1 | 22.4 | 25.3 | 26.4 | 28.1 | 29.0 | 31.8 | 32.6 | 33.0 | 37.0 | 18.3 |
| 2002 | 14.8 | 22.8 | 24.9 | 26.2 | 27.6 | 29.3 | 30.6 | 30.9 | 36.9 | 35.0 | 23.1 |

Table 6.4.2 1 Stratified mean catch ( $\mathrm{Kg} /$ haul and Number/haul) and standard error of BLUE WHITING in bottom trawl surveys in Spanish waters (Divisions VIIIc and IXa north). All surveys in September-October.

| $\mathrm{Kg} / \mathrm{haul}$ | $30-100 \mathrm{~m}$ |  | $101-200 \mathrm{~m}$ |  | $201-500 \mathrm{~m}$ |  | TOTAL $30-500 \mathrm{~m}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | Mean | SD | Mean | SD | Mean |  |
| 1985 | 9.50 | 5.87 | 119.75 | 45.99 | 68.18 | 13.79 | 92.83 | 28.24 |
| 1986 | 9.74 | 7.13 | 45.41 | 12.37 | 29.54 | 8.70 | 36.93 | 7.95 |
| 1987 | - | - | - | - | - | -93 | - |  |
| 1988 | 2.90 | 2.59 | 154.12 | 38.69 | 183.07 | 141.94 | 143.30 | 45.84 |
| 1989 | 14.17 | 12.03 | 76.92 | 17.08 | 18.79 | 6.23 | 59.00 | 11.68 |
| 1990 | 6.25 | 3.29 | 52.54 | 9.00 | 18.80 | 4.99 | 43.60 | 6.60 |
| 1991 | 64.59 | 34.65 | 126.41 | 26.06 | 46.07 | 18.99 | 97.10 | 17.16 |
| 1992 | 6.37 | 2.59 | 44.12 | 6.64 | 29.50 | 6.16 | 34.60 | 4.23 |
| 1993 | 1.06 | 0.63 | 14.07 | 3.73 | 51.08 | 22.02 | 22.59 | 6.44 |
| 1994 | 8.04 | 5.28 | 37.18 | 8.45 | 25.42 | 5.27 | 29.70 | 5.19 |
| 1995 | 19.97 | 13.87 | 36.43 | 4.82 | 15.97 | 4.10 | 28.52 | 3.66 |
| 1996 | 7.27 | 3.95 | 49.23 | 7.19 | 92.54 | 17.76 | 54.52 | 6.36 |
| 1997 | 7.60 | 4.44 | 44.21 | 10.61 | 60.18 | 17.54 | 44.01 | 8.00 |
| 1998 | 5.29 | 1.92 | 41.09 | 7.64 | 73.80 | 24.06 | 44.48 | 7.82 |
| 1999 | 31.41 | 7.28 | 108.46 | 17.24 | 150.24 | 39.53 | 108.12 | 14.62 |
| 2000 | 39.52 | 9.73 | 88.89 | 14.32 | 62.23 | 27.65 | 74.42 | 11.25 |


|  | $70-120 \mathrm{~m}$ |  | $121-200 \mathrm{~m}$ |  | $201-500 \mathrm{~m}$ | TOTAL $70-500 \mathrm{~m}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 17.87 | 7.35 | 44.68 | 10.52 | 57.14 | 16.60 | 42.62 |
| 1998 | 14.13 | 4.17 | 42.78 | 8.13 | 78.88 | 22.01 | 47.14 |
| 1999 | 92.66 | 14.60 | 111.76 | 19.87 | 169.21 | 50.26 | 124.27 |
| 2000 | 62.39 | 12.00 | 91.99 | 14.75 | 58.72 | 24.94 | 76.19 |
| 2001 | 8.35 | 3.31 | 50.18 | 10.09 | 52.41 | 16.71 | 4.58 |
| 2002 | 31.40 | 5.02 | 69.00 | 13.41 | 36.75 | 12.07 | 10.83 |


| Number/haul | 30-100 m |  | 101-200 m |  | 201-500 m |  | TOTAL 30-500 m |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | Mean | SD | Mean | SD | Mean | SD |
| 1985 | 267 | 181.71 | 3669 | 1578.86 | 1377 | 262.98 | 2644 | 963.20 |
| 1986 | 368 | 237.56 | 2486 | 1006.67 | 752 | 238.87 | 1763 | 616.40 |
| 1987 | - | - | - | - | - | - | - | - |
| 1988 | 83 | 71.74 | 6112 | 1847.36 | 7276 | 6339.88 | 5694 | 2086.00 |
| 1989 | 629 | 537.29 | 3197 | 876.75 | 566 | 213.11 | 2412 | 599.00 |
| 1990 | 220 | 115.48 | 2219 | 426.46 | 578 | 185.43 | 1722 | 276.00 |
| 1991 | 2922 | 1645.73 | 5563 | 1184.69 | 1789 | 847.33 | 4214 | 780.88 |
| 1992 | 124 | 50.81 | 1412 | 233.99 | 845 | 199.12 | 1069 | 146.87 |
| 1993 | 14 | 8.61 | 257 | 69.61 | 894 | 427.77 | 401 | 124.53 |
| 1994 | 346 | 234.12 | 2002 | 456.50 | 997 | 245.91 | 1487 | 689.00 |
| 1995 | 1291 | 864.97 | 2004 | 341.48 | 485 | 137.81 | 1493 | 240.37 |
| 1996 | 147 | 82.71 | 1167 | 167.20 | 2097 | 385.23 | 1263 | 142.30 |
| 1997 | 224 | 121.69 | 1425 | 359.12 | 1254 | 330.37 | 1228 | 234.50 |
| 1998 | 123 | 44.12 | 1442 | 334.24 | 1823 | 592.92 | 1347 | 251.37 |
| 1999 | 795 | 218.58 | 3996 | 697.66 | 5279 | 1521.62 | 3861 | 576.10 |
| 2000 | 1574 | 360.78 | 3701 | 568.17 | 2036 | 857.01 | 2940 | 406.62 |
|  | 70-120 m |  | 121-200 m |  | 201-500 m |  | TOTAL 70-500 m |  |
| 1997 | 552 | 235.60 | 1443 | 361.89 | 1183 | 323.14 | 1180 | 209.94 |
| 1998 | 351 | 105.96 | 1463 | 320.26 | 2012 | 590.04 | 1387 | 234.82 |
| 1999 | 2502 | 427.23 | 4358 | 847.87 | 6119 | 2026.39 | 4474 | 727.32 |
| 2000 | 2267 | 414.97 | 3930 | 604.11 | 2009 | 859.71 | 3027 | 400.87 |
| 2001 | 171 | 77.34 | 1310 | 263.84 | 1232 | 381.49 | 1048 | 172.74 |
| 2002 | 771 | 90.34 | 2526 | 499.30 | 1075 | 331.09 | 1739 | 268.70 |

Table 6.4.2 2 BLUE WHITING. Stratified mean catch ( $\mathrm{Kg} / \mathrm{haul}$ ) and standard error of in bottom trawl surveys in Portuguese waters (Division IXa ).

|  |  | 20-100 m |  | 100-200 m |  | $200-500 \mathrm{~m}$ |  | $500-750 \mathrm{~m}$ |  | TOTAL: 20-500 m |  |  | TOTAL: 20-750 m |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Month | y | sy | y | sy | y | sy | y | sy | y |  | sy | y |  | sy |
| 1979 | June | 0 | 0 | 33 | 23 | 86 | 35 | - | - |  | 31 | 12 |  | - | - |
|  | Oct./Nov. | 5 | 5 | 17 | 8 | 103 | 48 | - | - |  | 28 | 9 |  | - | - |
| 1980 | March | 0 | 0 | 178 | 173 | 5 | 1 | - | - |  | 72 | 69 |  | - | - |
|  | May/June | 1 | 3 | 4 | 2 | 45 | 18 | - | - |  | 11 | 4 |  | - | - |
|  | October | 4 | 3 | 10 | 4 | 587 | 306 | - | - |  | 117 | 58 |  | - | - |
| 1981 | March | 0 | 0 | 24 | 17 | 186 | 113 | - | - |  | 42 | 22 |  | - | - |
|  | June | 0 | 0 | 4 | 2 | 178 | 25 | - | - |  | 34 | 4 |  | - | - |
| 1982 | April/May | 0 | 0 | 3 | 3 | 136 | 39 | - | - |  | 26 | 7 |  | - | - |
|  | September | 1 | 1 | 85 | 42 | 271 | 123 | - | - |  | 86 | 29 |  | - | - |
| 1983 | March | 1 | 1 | 14 | 10 | 259 | 96 | - | - |  | 54 | 18 |  | - | - |
|  | June | 0 | 0 | 23 | 8 | 177 | 47 | - | - |  | 42 | 9 |  | - | - |
| 1985 | June | 0 | 0 | 194 | 146 | 405 | 162 | - | - |  | 159 | 68 |  | - | - |
|  | October | 4 | 3 | 133 | 84 | 341 | 39 | - | - |  | 120 | 35 |  | - | - |
| 1986 | June | 4 | 1 | 59 | 19 | 196 | 31 | - | - |  | 65 | 10 |  | - | - |
|  | October | 2 | 1 | 357 | 144 | 650 | 111 | - | - |  | 276 | 63 |  | - | - |
| 1987 | October | 3 | 0 | 297 | 64 | 747 | 229 | - | - |  | 263 | 50 |  | - | - |
| 1988 | October | 4 | 2 | 165 | 47 | 457 | 106 | - | - |  | 155 | 28 |  | - | - |
| 1989 | July | 0 | 0 | 42 | 21 | 323 | 143 | 79 | 36 |  | - | - |  | 78 | 24 |
|  | October | 7 | 4 | 70 | 26 | 306 | 84 | 24 | 2 |  | - | - |  | 79 | 16 |
| 1990 | July | 2 | 2 | 153 | 103 | 242 | 42 | 50 | 5 |  | - | - |  | 96 | 35 |
|  | October | 11 | 5 | 90 | 28 | 762 | 234 | 42 | 10 |  | - | - |  | 153 | 35 |
| 1991 | July | 1 | 1 | 140 | 40 | 268 | 38 | 64 | 18 |  | - | - |  | 98 | 15 |
|  | October | 8 | 5 | 83 | 18 | 259 | 53 | 121 | 27 |  | - | - |  | 91 | 11 |
| 1992 | February | 7 | 7 | 43 | 35 | 249 | 21 | 73 | 3 |  |  |  |  | 68 | 12 |
|  | July | 1 | 1 | 29 | 18 | 216 | 43 | 27 | 5 |  | - | - |  | 47 | 9 |
|  | October | 1 | 1 | 22 | 7 | 208 | 44 | 80 | 3 |  | - | - |  | 54 | 7 |
| 1993 | February | 0 | 0 | 19 | 14 | 105 | 31 | 36 | 0 |  | - | - |  | 42 | 10 |
|  | July | 0 | 0 | 3 | 3 | 151 | 28 | 55 | 5 |  | - | - |  | 34 | 4 |
|  | November | 0 | 0 | 90 | 0 | 189 | 43 | 6 | 1 |  | - | - |  | 86 | 9 |
| 1994 | October | 0 | 0 | 374 | 30 | 283 | 32 | 49 | 7 |  | - | - |  | 174 | 11 |
| 1995 | July | 0 | 0 | 18 | 14 | 130 | 20 | 52 | 3 |  | - | - |  | 35 | 5 |
|  | October | 18 | 15 | 103 | 21 | 328 | 91 | 31 | 12 |  | - | - |  | 94 | 16 |
| 1996 | October | 25 | 24 | 12 | 2 | 36 | 6 | 25 | 7 |  |  |  |  | 22 | 8 |
| 1997 | June | 0 | 0 | 3 | 3 | 116 | 42 | 45 | 12 |  | - | - |  | 27 | 7 |
|  | October | 2 | 1 | 54 | 20 | 77 | 13 | 7 | 2 |  | - | - |  | 32 | 8 |
| 1998 | July | 0 | 0 | 8 | 5 | 105 | 17 | 38 | 3 |  | - | - |  | 25 | 3 |
|  | October | 1 | 1 | 384 | 87 | 427 | 101 | 20 | 2 |  | - | - |  | 212 | 36 |
| 1999 | July | 1 |  | 60 |  | 66 |  | 25 |  |  | - | - |  | 37 | $\mathrm{n} / \mathrm{a}$ |
|  | October | 0 |  | 70 |  | 78 |  | 18 |  |  | - | - |  | 41 | n/a |
| 2000 | July | 23 | 13 | 109 | 34 | 116 | 10 | 63 | 6 |  |  |  |  | 75 | 13 |
|  | October | 11 | 4 | 155 | 53 | 196 | 22 | 54 | 4 |  |  |  |  | 100 | 19 |
| 2001 | July | 18 | 7 | 238 | 37 | 305 | 37 | 57 | 14 |  |  |  |  | 152 | 23 |
|  | October | 106 | 6 | 474 | 224 | 294 | 66 | 0 | 0 |  |  |  |  | 295 | 97 |
| 2002 | July | n/a |  | n/a |  | n/a |  | $\mathrm{n} / \mathrm{a}$ |  |  |  |  |  | n/a |  |
|  | October | 19 | 12 | 176 | 81 | 180 | 24 | 0 | 0 |  |  |  |  | 116 | 34 |

Table 6.4.3.1. Age stratified Spanish cpue (not used in the assessment)
units

| Numbers | age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ | total |
| 1981 |  |  |  |  |  |  |  |  |  |  |
| 1982 |  |  |  |  |  |  |  |  |  |  |
| 1983 |  | 7196 | 16392 | 9311 | 7476 | 6326 | 1718 |  |  |  |
| 1984 |  | 13710 | 27286 | 14845 | 4836 | 1755 | 1750 |  |  |  |
| 1985 |  | 14573 | 23823 | 14126 | 6256 | 1232 | 217 |  |  |  |
| 1986 |  | 3721 | 14131 | 14745 | 7113 | 1278 | 505 |  |  |  |
| 1987 |  | 25328 | 13153 | 6664 | 2938 | 1029 | 166 |  |  |  |
| 1988 |  | 7778 | 21473 | 18436 | 6391 | 1300 | 781 |  |  |  |
| 1989 |  | 15272 | 18486 | 17160 | 8374 | 3760 | 1003 |  |  |  |
| 1990 |  | 21444 | 19407 | 5194 | 1803 | 1357 | 451 |  |  |  |
| 1991 |  | 15924 | 15370 | 4989 | 2329 | 1045 | 440 |  |  |  |
| 1992 |  | 10007 | 24235 | 9671 | 4316 | 1194 | 462 |  |  |  |
| 1993 |  | 4036 | 13991 | 22493 | 7979 | 1354 | 658 |  |  |  |
| 1994 |  | 543 | 6066 | 15917 | 7474 | 2990 | 1055 |  |  |  |
| 1995 |  | 9090 | 14409 | 6833 | 4551 | 1990 | 623 |  |  |  |
| 1996 |  | 3905 | 14557 | 14449 | 3931 | 3639 | 1834 |  |  |  |
| 1997 |  | 8742 | 15875 | 11134 | 3698 | 1046 | 450 |  |  |  |
| 1998 |  | 5884 | 13236 | 9803 | 10844 | 5229 | 1153 |  |  |  |
| 1999 |  | 2048 | 10268 | 20242 | 9833 | 6287 | 3047 |  |  |  |
| 2000 |  | 6207 | 15518 | 13987 | 5375 | 1264 | 1414 |  |  |  |
| 2001 |  | 16223 | 16488 | 6830 | 1620 | 1148 | 162 |  |  |  |
| 2002 |  | 10520 | 13725 | 10265 | 3385 | 336 | 69 |  |  |  |

Table 6.4.4.4.1 Blue Whiting. ICA assessment2003. Stock summary table

| Year | Recruits thousands | Total Biomass tonnes | Spawning Biomass tonnes | Landings <br> tonnes | Yield /SSB ratio | Mean F Ages <br> 3-7 | $\begin{aligned} & \hline \text { SoP } \\ & (\%) \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 3459600 | 3644309 | 2724894 | 907732 | 0.3331 | 0.267 | 98 |
| 1982 | 4386890 | 2818864 | 2097158 | 513203 | 0.2447 | 0.2044 | 93 |
| 1983 | 16725260 | 2748712 | 1570613 | 561332 | 0.3574 | 0.2378 | 101 |
| 1984 | 19151220 | 2852090 | 1411123 | 626592 | 0.444 | 0.2916 | 101 |
| 1985 | 10416520 | 3053277 | 1677750 | 676812 | 0.4034 | 0.351 | 99 |
| 1986 | 8198360 | 3174858 | 1944551 | 801786 | 0.4123 | 0.5108 | 94 |
| 1987 | 8538260 | 2846074 | 1671290 | 656588 | 0.3929 | 0.4292 | 100 |
| 1988 | 6742980 | 2443886 | 1422149 | 552020 | 0.3882 | 0.5224 | 99 |
| 1989 | 8901480 | 2458121 | 1333307 | 598147 | 0.4486 | 0.5575 | 94 |
| 1990 | 20535850 | 2697373 | 1247517 | 558788 | 0.4479 | 0.5362 | 100 |
| 1991 | 8749600 | 3272796 | 1671266 | 363724 | 0.2176 | 0.2831 | 99 |
| 1992 | 5889070 | 3423956 | 2208659 | 473789 | 0.2145 | 0.1998 | 99 |
| 1993 | 5401120 | 3179116 | 2117282 | 475143 | 0.2244 | 0.2188 | 99 |
| 1994 | 6256080 | 3049748 | 2038531 | 458028 | 0.2247 | 0.2165 | 100 |
| 1995 | 8502100 | 3007156 | 1843697 | 505938 | 0.2744 | 0.272 | 100 |
| 1996 | 21676420 | 3201244 | 1701938 | 629286 | 0.3697 | 0.3358 | 101 |
| 1997 | 45724530 | 4933363 | 1926080 | 640089 | 0.3323 | 0.3132 | 100 |
| 1998 | 26356140 | 6033235 | 2703874 | 1123732 | 0.4156 | 0.5046 | 99 |
| 1999 | 16860020 | 6260805 | 3379582 | 1251463 | 0.3703 | 0.4123 | 99 |
| 2000 | 36414130 | 6850662 | 3311202 | 1409143 | 0.4256 | 0.5284 | 99 |
| 2001 | 56793570 | 8223035 | 3434237 | 1775305 | 0.5169 | 0.5535 | 100 |
| 2002 | 27291060 | 8454036 | 4040006 | 1556955 | 0.3854 | 0.6216 | 100 |

Table 6.4.5.1.1 Tuning data for the blue whiting assessment. Inside the framed areas constant selection pattern is assumed. $-1=$ missing data.


Norwegian acoustic spawning stock survey, ages 2-8

|  |  |  |  | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 1 | 1 | 1 | 2372 | 7583 | 3253 | 3647 | 4611 | 4638 | 3654 |
| 1982 | 1 | 1 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1983 | 1 | 1 | 1 | 297 | 2108 | 2723 | 6511 | 3735 | 3650 | 3153 |
| 1984 | 1 | 1 | 1 | 15767 | 1721 | 1616 | 1719 | 1858 | 1128 | 567 |
| 1985 | 1 | 1 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1986 | 1 | 1 | 1 | 1003 | 5829 | 4122 | 624 | 228 | 203 | 250 |
| 1987 | 1 | 1 | 1 | 4960 | 8417 | 22589 | 4735 | 282 | 417 | 385 |
| 1988 | 1 | 1 | 1 | 9712 | 9090 | 12367 | 20392 | 7355 | 723 | 599 |
| 1989 | 1 | 1 | 1 | 6787 | 22270 | 9973 | 10504 | 7803 | 933 | 293 |
| 1990 | 1 | 1 | 1 | 14169 | 12670 | 11228 | 5587 | 6556 | 3273 | 516 |
| 1991 | 1 | 1 | 1 | 11147 | 6340 | 8497 | 7407 | 4558 | 2019 | 545 |
| 1992 | 1 | 1 | 1 | 1232 | 26123 | 4719 | 1574 | 1386 | 810 | 616 |
| 1993 | 1 | 1 | 1 | 4489 | 3321 | 26771 | 2643 | 1270 | 557 | 426 |
| 1994 | 1 | 1 | 1 | 1603 | 2950 | 4476 | 11354 | 1742 | 1687 | 908 |
| 1995 | 1 | 1 | 1 | 8538 | 9874 | 7906 | 6861 | 9467 | 1795 | 1083 |
| 1996 | 1 | 1 | 1 | 8781 | 7433 | 8371 | 2399 | 4455 | 4111 | 1202 |
| 1997 | 1 | 1 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1998 | 1 | 1 | 1 | 18218 | 34991 | 4697 | 1674 | 279 | 407 | 381 |
| 1999 | 1 | 1 | 1 | 19034 | 60309 | 26103 | 1481 | 316 | 72 | 153 |
| 2000 | 1 | 1 | 1 | 8613 | 31011 | 41382 | 6843 | 898 | 427 | 228 |
| 2001 | 1 | 1 | 1 | 44162 | 12843 | 13805 | 8292 | 718 | 175 | 51 |
| 2002 | 1 | 1 | 1 | 71996 | 54740 | 12757 | 5266 | 8404 | 1450 | 305 |
| 2003 | 1 | 1 | 1 | 23992 | 70303 | 28756 | 5735 | 2430 | 1708 | 260 |

Russian acoustic spawning stock survey, ages 3-8

|  |  |  |  | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 1 | 2 | 10 | 540 | 2750 | 1340 | 1380 | 1570 | 2350 |
| 1983 | 1 | 2 | 10 | 2330 | 2930 | 9390 | 3880 | 1970 | 1370 |
| 1984 | 1 | 2 | 10 | 2900 | 800 | 1100 | 4200 | 2200 | 1200 |
| 1985 | 1 | 2 | 10 | 13220 | 930 | 580 | 1780 | 860 | 610 |
| 1986 | 1 | 2 | 10 | 18750 | 23180 | 2540 | 610 | 620 | 750 |
| 1987 | 1 | 2 | 10 | 4480 | 19170 | 5860 | 1070 | 500 | 810 |
| 1988 | 1 | 2 | 10 | 3710 | 4550 | 8610 | 4130 | 1270 | 480 |
| 1989 | 1 | 2 | 10 | 11910 | 7120 | 6670 | 6970 | 4580 | 2750 |
| 1990 | 1 | 2 | 10 | 9740 | 12140 | 5740 | 2580 | 1470 | 220 |
| 1991 | 1 | 2 | 10 | 10300 | 5350 | 5130 | 2630 | 1770 | 870 |
| 1992 | 1 | 2 | 1 | 20010 | 6700 | 1350 | 440 | 390 | 170 |
| 1993 | 1 | 2 | 1 | 4728 | 12337 | 5304 | 2249 | 1316 | 621 |
| 1994 | 1 | 2 | 1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1995 | 1 | 2 | 1 | 12657 | 10028 | 8942 | 2651 | 1093 | 408 |
| 1996 | 1 | 2 | 1 | 15285 | 10629 | 4897 | 6940 | 1482 | 653 |

Norwegian Sea acoustic survey, ages 1-7

|  |  |  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 3 | 3 | 1 | 182 | 728 | 4542 | 3874 | 2678 | 2834 | 2964 |
| 1982 | 3 | 3 | 1 | 184 | 460 | 1242 | 4715 | 3611 | 3128 | 2323 |
| 1983 | 3 | 3 | 1 | 22356 | 396 | 468 | 756 | 1404 | 576 | 468 |
| 1984 | 3 | 3 | 1 | 30380 | 13916 | 833 | 392 | 539 | 539 | 343 |
| 1985 | 3 | 3 | 1 | 5969 | 23876 | 12502 | 658 | 423 | 188 | 235 |
| 1986 | 3 | 3 | 1 | 2324 | 2380 | 7224 | 6944 | 1876 | 952 | 336 |
| 1987 | 3 | 3 | 1 | 8204 | 4032 | 5180 | 5572 | 1204 | 224 | 168 |
| 1988 | 3 | 3 | 1 | 4992 | 2880 | 2640 | 3480 | 912 | 120 | 96 |
| 1989 | 3 | 3 | 1 | 1172 | 1125 | 812 | 379 | 410 | 212 | 22 |
| 1990 | 3 | 3 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1991 | 3 | 3 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1992 | 3 | 3 | 1 | 792 | 1134 | 6939 | 766 | 247 | 172 | 90 |
| 1993 | 3 | 3 | 1 | 830 | 125 | 1070 | 6392 | 1222 | 489 | 248 |
| 1994 | 3 | 3 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1995 | 3 | 3 | 1 | 6974 | 2811 | 1999 | 1209 | 1622 | 775 | 173 |
| 1996 | 3 | 3 | 1 | 23464 | 1057 | 899 | 649 | 436 | 505 | 755 |
| 1997 | 3 | 3 | 1 | 30227 | 25638 | 1524 | 779 | 300 | 407 | 260 |
| 1998 | 3 | 3 | 1 | 24244 | 47815 | 16282 | 556 | 212 | 100 | 64 |
| 1999 | 3 | 3 | 1 | 14367 | 9750 | 23701 | 9754 | 1733 | 466 | 79 |
| 2000 | 3 | 3 | 1 | 25813 | 3298 | 2721 | 3078 | 23 | 46 | 6 |
| 2001 | 3 | 3 | 1 | 61470 | 22051 | 7883 | 3225 | 1824 | 156 | 12 |

Table 6.4.5.1.2 Blue Whiting, Total yearly fishing mortalities at age $\left(\mathrm{Fref}=\mathrm{F}_{3-7 \mathrm{u}}\right)$

Output from AMCI run id 20030504131214.494

|  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.0443 | 0.0377 | 0.0952 | 0.1304 | 0.1404 | 0.1660 | 0.1354 | 0.1063 |
| 2 | 0.0666 | 0.0602 | 0.0938 | 0.1485 | 0.1681 | 0.2155 | 0.1801 | 0.1550 |
| 3 | 0.1119 | 0.0951 | 0.1263 | 0.1751 | 0.2180 | 0.2891 | 0.2437 | 0.2096 |
| 4 | 0.1444 | 0.1240 | 0.1608 | 0.2121 | 0.2379 | 0.3465 | 0.3172 | 0.2749 |
| 5 | 0.1706 | 0.1375 | 0.1742 | 0.2308 | 0.2556 | 0.3581 | 0.3275 | 0.3061 |
| 6 | 0.2291 | 0.1835 | 0.2353 | 0.3122 | 0.3565 | 0.4700 | 0.4025 | 0.3808 |
| 7 | 0.2640 | 0.2095 | 0.2695 | 0.3503 | 0.3852 | 0.5145 | 0.4482 | 0.3908 |
| 8 | 0.3146 | 0.2508 | 0.3199 | 0.4093 | 0.4484 | 0.5898 | 0.5284 | 0.4602 |
| 9 | 0.3463 | 0.2763 | 0.3496 | 0.4423 | 0.4736 | 0.6440 | 0.5620 | 0.4812 |
| 10 | 0.3305 | 0.2636 | 0.3348 | 0.4258 | 0.4610 | 0.6169 | 0.5452 | 0.4707 |
| Fref | 0.1840 | 0.1499 | 0.1932 | 0.2561 | 0.2906 | 0.3956 | 0.3478 | 0.3124 |
|  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 1 | 0.1162 | 0.1036 | 0.0391 | 0.0472 | 0.0453 | 0.0461 | 0.0535 | 0.0758 |
| 2 | 0.1725 | 0.1594 | 0.0651 | 0.0711 | 0.0670 | 0.0623 | 0.0774 | 0.1024 |
| 3 | 0.2476 | 0.2311 | 0.0958 | 0.1046 | 0.1004 | 0.0996 | 0.1245 | 0.1692 |
| 4 | 0.3142 | 0.2990 | 0.1220 | 0.1329 | 0.1305 | 0.1260 | 0.1585 | 0.2153 |
| 5 | 0.3618 | 0.3463 | 0.1469 | 0.1511 | 0.1454 | 0.1482 | 0.1872 | 0.2431 |
| 6 | 0.4388 | 0.4365 | 0.1860 | 0.1891 | 0.1758 | 0.1728 | 0.2107 | 0.2804 |
| 7 | 0.4285 | 0.4055 | 0.1735 | 0.1819 | 0.1678 | 0.1687 | 0.2099 | 0.2750 |
| 8 | 0.5127 | 0.4696 | 0.1853 | 0.1955 | 0.1812 | 0.1770 | 0.2239 | 0.3004 |
| 9 | 0.5328 | 0.5017 | 0.1925 | 0.1916 | 0.1767 | 0.1833 | 0.2201 | 0.2898 |
| 10 | 0.5228 | 0.4856 | 0.1889 | 0.1936 | 0.1789 | 0.1802 | 0.2220 | 0.2951 |
| Fref | 0.3582 | 0.3437 | 0.1448 | 0.1519 | 0.1440 | 0.1431 | 0.1782 | 0.2366 |


|  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |
| 1 | 0.0701 | 0.0893 | 0.0750 | 0.0868 | 0.1178 | 0.1461 |
| 2 | 0.1017 | 0.1382 | 0.1218 | 0.1462 | 0.1786 | 0.1974 |
| 3 | 0.1746 | 0.2475 | 0.2283 | 0.2817 | 0.3398 | 0.3683 |
| 4 | 0.2201 | 0.3109 | 0.3063 | 0.3772 | 0.4549 | 0.5186 |
| 5 | 0.2391 | 0.3234 | 0.2912 | 0.3629 | 0.4167 | 0.4721 |
| 6 | 0.2749 | 0.3766 | 0.3344 | 0.4281 | 0.4765 | 0.5219 |
| 7 | 0.2720 | 0.3818 | 0.3436 | 0.4289 | 0.4874 | 0.5209 |
| 8 | 0.2928 | 0.4010 | 0.3705 | 0.4680 | 0.5349 | 0.6219 |
| 9 | 0.2833 | 0.3717 | 0.3233 | 0.3888 | 0.4692 | 0.5837 |
| 10 | 0.2881 | 0.3863 | 0.3469 | 0.4284 | 0.5021 | 0.6028 |
|  |  |  |  |  |  |  |
| Fref | 0.2362 | 0.3280 | 0.3008 | 0.3758 | 0.4351 | 0.4804 |

Table 6.4.5.1.3 Blue Whiting, Stock numbers at age at 1 Jan.
Output from AMCI run id 20030504131214.494

|  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 3640.3 | 3881.5 | 10579.6 | 18515.4 | 12451.7 | 10136.9 | 10296.5 | 8921.7 |
| 2 | 3885.0 | 2851.2 | 3060.2 | 7875.5 | 13305.6 | 8859.3 | 7030.1 | 7362.3 |
| 3 | 5026.8 | 2975.8 | 2198.1 | 2281.2 | 5558.2 | 9208.5 | 5847.2 | 4807.2 |
| 4 | 3645.6 | 3680.0 | 2215.3 | 1586.2 | 1567.7 | 3659.5 | 5646.5 | 3752.0 |
| 5 | 3177.3 | 2583.3 | 2661.5 | 1544.4 | 1050.5 | 1011.8 | 2118.8 | 3366.4 |
| 6 | 3213.8 | 2193.4 | 1843.4 | 1830.7 | 1003.8 | 666.1 | 579.1 | 1250.3 |
| 7 | 2744.6 | 2092.5 | 1494.7 | 1192.9 | 1096.9 | 575.3 | 340.9 | 317.0 |
| 8 | 2739.9 | 1725.6 | 1389.4 | 934.7 | 688.0 | 610.9 | 281.6 | 178.3 |
| 9 | 2225.7 | 1637.7 | 1099.4 | 826.1 | 508.2 | 359.7 | 277.3 | 135.9 |
| 10 | 7531.2 | 3279.8 | 1944.2 | 1146.3 | 710.1 | 423.9 | 238.8 | 180.4 |
|  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 1 | 10351.5 | 24666.4 | 10149.1 | 6660.7 | 5904.7 | 6308.9 | 8131.0 | 22987.8 |
| 2 | 6568.2 | 7545.7 | 18208.5 | 7990.4 | 5202.2 | 4620.2 | 4932.5 | 6310.3 |
| 3 | 5162.2 | 4525.8 | 5267.4 | 13968.8 | 6093.0 | 3983.2 | 3554.1 | 3737.5 |
| 4 | 3191.6 | 3299.5 | 2940.7 | 3918.8 | 10301.3 | 4511.9 | 2951.8 | 2569.3 |
| 5 | 2333.5 | 1908.5 | 2003.2 | 2131.1 | 2809.1 | 7401.9 | 3256.7 | 2062.5 |
| 6 | 2029.5 | 1330.5 | 1105.2 | 1416.0 | 1500.1 | 1988.6 | 5225.6 | 2211.3 |
| 7 | 699.5 | 1071.4 | 704.1 | 751.2 | 959.6 | 1030.2 | 1369.8 | 3465.4 |
| 8 | 175.6 | 373.1 | 584.8 | 484.6 | 512.8 | 664.3 | 712.5 | 909.1 |
| 9 | 92.1 | 86.1 | 191.0 | 397.8 | 326.3 | 350.2 | 455.7 | 466.3 |
| 10 | 110.2 | 68.3 | 58.1 | 146.7 | 313.3 | 320.3 | 337.1 | 398.7 |
|  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |  |
| 1 | 49221.4 | 27893.3 | 17104.9 | 30165.9 | 30000.0 | 11500.0 | 11500.0 |  |
| 2 | 17446.6 | 37572.6 | 20885.1 | 12993.0 | 22643.9 | 21831.6 | 8135.4 |  |
| 3 | 4663.7 | 12902.8 | 26791.5 | 15138.9 | 9190.5 | 15506.5 | 14671.6 |  |
| 4 | 2583.7 | 3206.4 | 8247.9 | 17457.1 | 9351.8 | 5357.0 | 8784.3 |  |
| 5 | 1696.1 | 1697.4 | 1923.7 | 4971.0 | 9801.8 | 4858.0 | 2611.3 |  |
| 6 | 1324.2 | 1093.3 | 1005.8 | 1177.1 | 2831.3 | 5290.3 | 2480.6 |  |
| 7 | 1367.8 | 823.6 | 614.2 | 589.4 | 628.1 | 1439.3 | 2570.2 |  |
| 8 | 2155.1 | 853.1 | 460.3 | 356.7 | 314.2 | 315.8 | 699.9 |  |
| 9 | 551.2 | 1316.6 | 467.8 | 260.2 | 182.9 | 150.7 | 138.8 |  |
| 10 | 394.9 | 448.9 | 855.5 | 499.6 | 264.2 | 152.5 | 99.5 |  |

Table 6.4.5.1.4 Blue Whiting, Stock summary
Output from AMCI run id 20030504131214.494

| Year | Recruits <br> age | SSB | F | Catch <br> SOP |
| :--- | ---: | ---: | ---: | ---: |
| 1981 | 3640338 | 4333367 | 0.1840 | 922980 |
| 1982 | 3881467 | 3142812 | 0.1499 | 550643 |
| 1983 | 10579612 | 2335954 | 0.1932 | 553344 |
| 1984 | 18515411 | 1772338 | 0.2561 | 615569 |
| 1985 | 12451709 | 1841415 | 0.2906 | 678214 |
| 1986 | 10136893 | 2072225 | 0.3956 | 847145 |
| 1987 | 10296488 | 1883688 | 0.3478 | 654718 |
| 1988 | 8921724 | 1726283 | 0.3124 | 552264 |
| 1989 | 10351479 | 1729522 | 0.3582 | 630316 |
| 1990 | 24666414 | 1707475 | 0.3437 | 558128 |
| 1991 | 10149052 | 2277819 | 0.1448 | 364008 |
| 1992 | 6660742 | 3012788 | 0.1519 | 474592 |
| 1993 | 5904686 | 2949885 | 0.1440 | 475198 |
| 1994 | 6308929 | 2916378 | 0.1431 | 457696 |
| 1995 | 8130963 | 2677761 | 0.1782 | 505175 |
| 1996 | 22987755 | 2497657 | 0.2366 | 621104 |
| 1997 | 49221383 | 2515304 | 0.2362 | 639680 |
| 1998 | 27893318 | 3441203 | 0.3280 | 1131954 |
| 1999 | 17104882 | 4093265 | 0.3008 | 1261033 |
| 2000 | 30165937 | 3987962 | 0.3758 | 1412449 |
| 2001 | 3000000 | 3943532 | 0.4351 | 1771805 |
| 2002 | 11500000 | 3789156 | 0.4804 | 1556954 |
| 2003 | 11500000 | 3283641 | 0.4804 | 0 |
| $2004 *$ | 11500000 | 2563313 | 0.4804 | 0 |
| $2005 *$ | 11500000 | 2030436 | 0.4804 | 0 |

Table 6.4.5.2.1
ISVPA, effort-controlled, restriction of zero sums for residuals in separable representation of $F$ minimization of SSE for catch-at-age and age-structured indexes.

SSE from catch-at-age and all age-structured indexes are taken with equal weights

|  | $\mathbf{R ( 1 )}$ | $\mathbf{B}(\mathbf{1 + )}$ | $\mathbf{S S B}$ | $\mathbf{F ( 3 - 7 )}$ |
| ---: | ---: | ---: | ---: | ---: |
| 1981 | 6751.3 | 4830.5 | 4076.3 | 0.240 |
| 1982 | 7956.2 | 4035.9 | 3261.0 | 0.177 |
| 1983 | 4631.7 | 3126.7 | 2457.1 | 0.302 |
| 1984 | 18575.1 | 2790.6 | 1886.4 | 0.363 |
| 1985 | 16584.4 | 3194.0 | 1899.1 | 0.386 |
| 1986 | 20492.0 | 3927.2 | 2405.9 | 0.525 |
| 1987 | 9446.7 | 3834.8 | 2393.4 | 0.514 |
| 1888 | 10465.6 | 3295.7 | 2186.7 | 0.580 |
| 1989 | 11085.9 | 3101.1 | 1935.3 | 0.533 |
| 1990 | 15622.4 | 2923.7 | 1771.2 | 0.609 |
| 1991 | 7320.9 | 2995.3 | 1843.2 | 0.238 |
| 1992 | 6586.1 | 3109.2 | 2209.9 | 0.214 |
| 1993 | 11116.6 | 3310.9 | 2163.7 | 0.193 |
| 1994 | 14490.8 | 3791.5 | 2343.7 | 0.206 |
| 1995 | 12257.2 | 4137.2 | 2579.2 | 0.249 |
| 1996 | 16891.5 | 3931.1 | 2602.1 | 0.301 |
| 1997 | 44517.3 | 5447.2 | 2768.0 | 0.274 |
| 1998 | 37028.6 | 6879.0 | 3529.6 | 0.400 |
| 1999 | 29224.7 | 7796.4 | 4357.7 | 0.329 |
| 2000 | 37988.7 | 8276.3 | 4771.5 | 0.337 |
| 2001 | 43059.4 | 9202.2 | 5362.1 | 0.292 |
| 2002 | 25954.0 | 9458.4 | 6095.2 | 0.296 |

Table 6.4.5.2.2
Abundance at age by years
(Blue whiting, ISVPA, based on signals from catch-at-age and all age-structured indexes (except Spanish CPUE))

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1981 | 6751.27 | 4112.83 | 7575.33 | 5596.75 | 2008.57 | 1880.31 | 1702.49 | 2316.71 | 2007.97 | 5072.04 |
| 1982 | 7956.18 | 5170.12 | 3059.21 | 5295.10 | 3693.08 | 1306.09 | 1143.98 | 1033.08 | 1293.35 | 3064.88 |
| 1983 | 4631.71 | 6192.72 | 3937.50 | 2225.24 | 3693.54 | 2549.01 | 859.77 | 751.67 | 640.55 | 950.86 |
| 1984 | 18575.07 | 3492.96 | 4504.10 | 2648.43 | 1390.48 | 2264.68 | 1432.72 | 481.58 | 376.51 | 623.44 |
| 1985 | 16584.36 | 13810.92 | 2488.00 | 2921.32 | 1570.49 | 805.56 | 1177.33 | 741.57 | 216.05 | 601.69 |
| 1986 | 20491.95 | 12265.43 | 9760.36 | 1591.51 | 1697.75 | 890.15 | 406.17 | 590.80 | 318.31 | 728.09 |
| 1987 | 9446.68 | 14717.61 | 8297.31 | 5771.77 | 822.87 | 846.84 | 373.25 | 169.08 | 191.49 | 362.53 |
| 1988 | 10465.61 | 6799.93 | 9989.64 | 4936.77 | 3012.07 | 414.66 | 360.50 | 157.78 | 56.14 | 85.06 |
| 1989 | 11085.89 | 7437.28 | 4527.11 | 5736.17 | 2440.25 | 1429.80 | 161.39 | 139.13 | 45.26 | 68.79 |
| 1990 | 15622.42 | 7950.30 | 5020.01 | 2666.24 | 2947.57 | 1208.96 | 593.51 | 66.50 | 44.37 | 156.12 |
| 1991 | 7320.89 | 11042.46 | 5250.17 | 2839.34 | 1287.48 | 1363.48 | 452.26 | 220.02 | 17.83 | 33.56 |
| 1992 | 6586.11 | 5608.76 | 8218.81 | 3673.77 | 1876.40 | 838.55 | 831.34 | 275.03 | 123.19 | 45.04 |
| 1993 | 11116.59 | 5077.66 | 4212.93 | 5840.94 | 2481.15 | 1251.00 | 527.39 | 521.65 | 160.53 | 73.74 |
| 1994 | 14490.79 | 8616.36 | 3843.64 | 3033.34 | 4017.08 | 1686.75 | 807.29 | 339.64 | 315.07 | 122.04 |
| 1995 | 12257.20 | 11195.94 | 6492.26 | 2746.02 | 2063.72 | 2699.40 | 1071.96 | 511.92 | 201.02 | 140.74 |
| 1996 | 16891.46 | 9365.96 | 8301.16 | 4513.36 | 1798.14 | 1330.91 | 1624.29 | 643.26 | 281.56 | 236.06 |
| 1997 | 44517.26 | 12744.05 | 6816.40 | 5589.64 | 2824.68 | 1104.39 | 749.79 | 911.92 | 323.24 | 524.37 |
| 1998 | 37028.55 | 33806.29 | 9363.84 | 4665.89 | 3580.98 | 1779.21 | 643.82 | 435.76 | 480.37 | 193.28 |
| 1999 | 29224.73 | 27304.37 | 23786.52 | 5943.41 | 2680.94 | 2004.72 | 881.59 | 317.43 | 182.35 | 372.38 |
| 2000 | 37988.65 | 21899.58 | 19674.00 | 15741.53 | 3628.41 | 1602.94 | 1088.64 | 476.89 | 151.52 | 363.33 |
| 2001 | 43059.41 | 28414.68 | 15737.07 | 12958.83 | 9545.23 | 2153.58 | 861.71 | 582.91 | 224.41 | 194.51 |
| 2002 | 25954.02 | 32550.72 | 20739.36 | 10649.42 | 8167.91 | 5907.58 | 1225.94 | 488.91 | 297.17 | 413.31 |

## Table 6.4.5.2.3

Fishing mortality coefficients by age by years
(Blue whiting, ISVPA, based on signals from catch-at-age and all age-structured indexes (except Spanish CPUE))

|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1981 | 0.0668 | 0.0960 | 0.1581 | 0.2157 | 0.2304 | 0.2969 | 0.2996 | 0.3829 | 0.3829 | 0.3829 |
| 1982 | 0.0506 | 0.0723 | 0.1183 | 0.1602 | 0.1708 | 0.2181 | 0.2200 | 0.2780 | 0.2780 | 0.2780 |
| 1983 | 0.0822 | 0.1184 | 0.1966 | 0.2702 | 0.2892 | 0.3761 | 0.3796 | 0.4914 | 0.4914 | 0.4914 |
| 1984 | 0.0964 | 0.1393 | 0.2330 | 0.3226 | 0.3459 | 0.4542 | 0.4586 | 0.6015 | 0.6015 | 0.6015 |
| 1985 | 0.1017 | 0.1471 | 0.2468 | 0.3427 | 0.3678 | 0.4848 | 0.4895 | 0.6457 | 0.6457 | 0.6457 |
| 1986 | 0.1310 | 0.1909 | 0.3254 | 0.4596 | 0.4956 | 0.6691 | 0.6764 | 0.9267 | 0.9267 | 0.9267 |
| 1987 | 0.1288 | 0.1875 | 0.3192 | 0.4504 | 0.4853 | 0.6540 | 0.6611 | 0.9025 | 0.9025 | 0.9025 |
| 1988 | 0.1416 | 0.2068 | 0.3548 | 0.5046 | 0.5451 | 0.7436 | 0.7520 | 1.0488 | 1.0488 | 1.0488 |
| 1989 | 0.1325 | 0.1931 | 0.3294 | 0.4658 | 0.5023 | 0.6792 | 0.6866 | 0.9429 | 0.9429 | 0.9429 |
| 1990 | 0.1470 | 0.2149 | 0.3699 | 0.5280 | 0.5709 | 0.7833 | 0.7923 | 1.1163 | 1.1163 | 1.1163 |
| 1991 | 0.0664 | 0.0953 | 0.1570 | 0.2142 | 0.2288 | 0.2948 | 0.2974 | 0.3800 | 0.3800 | 0.3800 |
| 1992 | 0.0601 | 0.0862 | 0.1415 | 0.1925 | 0.2054 | 0.2637 | 0.2660 | 0.3384 | 0.3384 | 0.3384 |
| 1993 | 0.0548 | 0.0784 | 0.1285 | 0.1743 | 0.1859 | 0.2380 | 0.2401 | 0.3042 | 0.3042 | 0.3042 |
| 1994 | 0.0580 | 0.0831 | 0.1363 | 0.1852 | 0.1975 | 0.2533 | 0.2555 | 0.3245 | 0.3245 | 0.3245 |
| 1995 | 0.0690 | 0.0992 | 0.1636 | 0.2234 | 0.2386 | 0.3080 | 0.3107 | 0.3978 | 0.3978 | 0.3978 |
| 1996 | 0.0817 | 0.1178 | 0.1955 | 0.2686 | 0.2875 | 0.3738 | 0.3773 | 0.4881 | 0.4881 | 0.4881 |
| 1997 | 0.0752 | 0.1082 | 0.1791 | 0.2453 | 0.2622 | 0.3396 | 0.3427 | 0.4410 | 0.4410 | 0.4410 |
| 1998 | 0.1046 | 0.1515 | 0.2546 | 0.3541 | 0.3801 | 0.5022 | 0.5072 | 0.6712 | 0.6712 | 0.6712 |
| 1999 | 0.0885 | 0.1277 | 0.2128 | 0.2935 | 0.3143 | 0.4106 | 0.4144 | 0.5395 | 0.5395 | 0.5395 |
| 2000 | 0.0904 | 0.1304 | 0.2175 | 0.3003 | 0.3217 | 0.4207 | 0.4246 | 0.5538 | 0.5538 | 0.5538 |
| 2001 | 0.0798 | 0.1149 | 0.1905 | 0.2616 | 0.2798 | 0.3634 | 0.3668 | 0.4737 | 0.4737 | 0.4737 |
| 2002 | 0.0807 | 0.1162 | 0.1929 | 0.2649 | 0.2834 | 0.3684 | 0.3717 | 0.4806 | 0.4806 | 0.4806 |

Table 6.5.1. Blue Whiting. Input data for the deterministic short-term prediction

## MFDP version 1a

Run: AMCI based Short term pred BW 2003
Time and date: 19:20 05/05/03
Fbar age range: 3-7

| 2003 |  |  |  |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Stock <br> size |  | Natural <br> mortality |  | Maturity <br> ogive | Prop. of F <br> bef. spaw. | Prop. of M <br> bef. spaw. | Weight <br> in stock |
| 1 | 12000 | 0.2 | 0.11 | 0.25 | 0.25 | 0.054 | Exploit. <br> pattern | Weight <br> in catch |
| 2 | 8135 | 0.2 | 0.40 | 0.25 | 0.25 | 0.076 | 0.177 | 0.054 |
| 3 | 14672 | 0.2 | 0.82 | 0.25 | 0.25 | 0.091 | 0.330 | 0.076 |
| 4 | 8784 | 0.2 | 0.86 | 0.25 | 0.25 | 0.109 | 0.450 | 0.109 |
| 5 | 2611 | 0.2 | 0.91 | 0.25 | 0.25 | 0.131 | 0.417 | 0.131 |
| 6 | 2481 | 0.2 | 0.94 | 0.25 | 0.25 | 0.158 | 0.476 | 0.158 |
| 7 | 2570 | 0.2 | 1.00 | 0.25 | 0.25 | 0.179 | 0.479 | 0.179 |
| 8 | 700 | 0.2 | 1.00 | 0.25 | 0.25 | 0.204 | 0.542 | 0.204 |
| 9 | 139 | 0.2 | 1.00 | 0.25 | 0.25 | 0.229 | 0.481 | 0.229 |
| 10 | 100 | 0.2 | 1.00 | 0.25 | 0.25 | 0.249 | 0.511 | 0.249 |


| 2004 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Stock size | Natural mortality | Maturity ogive | Prop. of F bef. spaw. | Prop. of M bef. spaw. | Weight in stock | Exploit. pattern | Weight in catch |
| 1 | 12000 | 0.2 | 0.11 | 0.25 | 0.25 | 0.054 | 0.117 | 0.054 |
| 2 | . | 0.2 | 0.40 | 0.25 | 0.25 | 0.076 | 0.174 | 0.076 |
| 3 | . | 0.2 | 0.82 | 0.25 | 0.25 | 0.091 | 0.330 | 0.091 |
| 4 | . | 0.2 | 0.86 | 0.25 | 0.25 | 0.109 | 0.450 | 0.109 |
| 5 | . | 0.2 | 0.91 | 0.25 | 0.25 | 0.131 | 0.417 | 0.131 |
| 6 | . | 0.2 | 0.94 | 0.25 | 0.25 | 0.158 | 0.476 | 0.158 |
| 7 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.179 | 0.479 | 0.179 |
| 8 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.204 | 0.542 | 0.204 |
| 9 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.229 | 0.481 | 0.229 |
| 10 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.249 | 0.511 | 0.249 |


| 2005 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Stock size | Natural mortality | Maturity ogive | Prop. of F bef. spaw. | Prop. of M bef. spaw. | Weight in stock | Exploit. pattern | Weight in catch |
| 1 | 12000 | 0.2 | 0.11 | 0.25 | 0.25 | 0.054 | 0.117 | 0.054 |
| 2 | . | 0.2 | 0.40 | 0.25 | 0.25 | 0.076 | 0.174 | 0.076 |
| 3 | . | 0.2 | 0.82 | 0.25 | 0.25 | 0.091 | 0.330 | 0.091 |
| 4 | . | 0.2 | 0.86 | 0.25 | 0.25 | 0.109 | 0.450 | 0.109 |
| 5 | . | 0.2 | 0.91 | 0.25 | 0.25 | 0.131 | 0.417 | 0.131 |
| 6 | . | 0.2 | 0.94 | 0.25 | 0.25 | 0.158 | 0.476 | 0.158 |
| 7 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.179 | 0.479 | 0.179 |
| 8 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.204 | 0.542 | 0.204 |
| 9 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.229 | 0.481 | 0.229 |
| 10 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.249 | 0.511 | 0.249 |

Input units are millions and kg - output in kilotonnes

Table 6.5.2. Blue Whiting. Input data for the deterministic short-term prediction

## MFDP version 1a

Run: ISVPA based Short term pred BW 2003
Time and date: 19:20 05/05/03
Fbar age range: 3-7

| 2003 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Stock size | Natural mortality | Maturity ogive | Prop. of F bef. spaw. | Prop. of M bef. spaw. | Weight in stock | Exploit. pattern | Weight in catch |
| 1 | 15000 | 0.2 | 0.11 | 0.25 | 0.25 | 0.054 | 0.084 | 0.054 |
| 2 | 19602 | 0.2 | 0.40 | 0.25 | 0.25 | 0.076 | 0.120 | 0.076 |
| 3 | 23726 | 0.2 | 0.82 | 0.25 | 0.25 | 0.091 | 0.201 | 0.091 |
| 4 | 14001 | 0.2 | 0.86 | 0.25 | 0.25 | 0.109 | 0.276 | 0.109 |
| 5 | 6690 | 0.2 | 0.91 | 0.25 | 0.25 | 0.131 | 0.295 | 0.131 |
| 6 | 5037 | 0.2 | 0.94 | 0.25 | 0.25 | 0.158 | 0.384 | 0.158 |
| 7 | 3346 | 0.2 | 1.00 | 0.25 | 0.25 | 0.179 | 0.388 | 0.179 |
| 8 | 692 | 0.2 | 1.00 | 0.25 | 0.25 | 0.204 | 0.503 | 0.204 |
| 9 | 248 | 0.2 | 1.00 | 0.25 | 0.25 | 0.229 | 0.503 | 0.229 |
| 10 | 360 | 0.2 | 1.00 | 0.25 | 0.25 | 0.249 | 0.503 | 0.249 |


| 2004 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Stock size | Natural mortality | Maturity ogive | Prop. of F bef. spaw. | Prop. of M bef. spaw. | Weight in stock | Exploit. pattern | Weight in catch |
| 1 | 15000 | 0.2 | 0.11 | 0.25 | 0.25 | 0.054 | 0.084 | 0.054 |
| 2 | . | 0.2 | 0.40 | 0.25 | 0.25 | 0.076 | 0.120 | 0.076 |
| 3 | . | 0.2 | 0.82 | 0.25 | 0.25 | 0.091 | 0.201 | 0.091 |
| 4 | . | 0.2 | 0.86 | 0.25 | 0.25 | 0.109 | 0.276 | 0.109 |
| 5 | . | 0.2 | 0.91 | 0.25 | 0.25 | 0.131 | 0.295 | 0.131 |
| 6 | . | 0.2 | 0.94 | 0.25 | 0.25 | 0.158 | 0.384 | 0.158 |
| 7 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.179 | 0.388 | 0.179 |
| 8 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.204 | 0.503 | 0.204 |
| 9 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.229 | 0.503 | 0.229 |
| 10 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.249 | 0.503 | 0.249 |


| 2005 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Stock size | Natural mortality | Maturity ogive | Prop. of F bef. spaw. | Prop. of M bef. spaw. | Weight in stock | Exploit. pattern | Weight in catch |
| 1 | 15000 | 0.2 | 0.11 | 0.25 | 0.25 | 0.054 | 0.084 | 0.054 |
| 2 | . | 0.2 | 0.40 | 0.25 | 0.25 | 0.076 | 0.120 | 0.076 |
| 3 | . | 0.2 | 0.82 | 0.25 | 0.25 | 0.091 | 0.201 | 0.091 |
| 4 | . | 0.2 | 0.86 | 0.25 | 0.25 | 0.109 | 0.276 | 0.109 |
| 5 | . | 0.2 | 0.91 | 0.25 | 0.25 | 0.131 | 0.295 | 0.131 |
| 6 | . | 0.2 | 0.94 | 0.25 | 0.25 | 0.158 | 0.384 | 0.158 |
| 7 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.179 | 0.388 | 0.179 |
| 8 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.204 | 0.503 | 0.204 |
| 9 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.229 | 0.503 | 0.229 |
| 10 | . | 0.2 | 1.00 | 0.25 | 0.25 | 0.249 | 0.503 | 0.249 |

Input units are millions and kg - output in kilotonnes
Table 6.5.3 Blue Whiting. Prediction with management option table. Basis for 2003: F2003 = F2002; Recruitment: GM 1981-2001 $=12000$ millions

## MFI based Short term pred BW 2003 Blue whiting combined stock, 2003 WG <br> Time and date: 19:20 05/05/03 <br> Fbar age range: 3-7

Basis for 2003: F2003 $=F(2000-2002)=0.43$; Recruitment: GM 1981-2001 $=12000$ millions

| 2003 |  |  |  |  | 2004 |  |  |  |  | 2005 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 4948 | 3083 | 1 | 0.4305 | 1290 | 4179 | 2776 | 0.0 | 0.000 | 0 | 4814 | 3319 |
|  |  |  |  |  | . | 2748 | 0.1 | 0.043 | 128 | 4676 | 3169 |
|  |  |  |  |  | . | 2720 | 0.2 | 0.086 | 252 | 4544 | 3027 |
|  |  |  |  |  | . | 2693 | 0.3 | 0.129 | 370 | 4416 | 2892 |
|  |  |  |  |  | . | 2665 | 0.4 | 0.172 | 485 | 4294 | 2763 |
|  |  |  |  |  |  | 2638 | 0.5 | 0.215 | 595 | 4176 | 2642 |
|  |  |  |  |  | . | 2612 | 0.6 | 0.258 | 700 | 4063 | 2526 |
|  |  |  |  |  | . | 2585 | 0.7 | 0.301 | 802 | 3954 | 2416 |
|  |  |  |  |  | . | 2559 | 0.8 | 0.344 | 900 | 3850 | 2312 |
|  |  |  |  |  | . | 2533 | 0.9 | 0.387 | 994 | 3749 | 2212 |
|  |  |  |  |  | . | 2508 | 1.0 | 0.431 | 1085 | 3652 | 2118 |
|  |  |  |  |  | . | 2483 | 1.1 | 0.474 | 1173 | 3559 | 2028 |
|  |  |  |  |  | . | 2458 | 1.2 | 0.517 | 1257 | 3470 | 1943 |
|  |  |  |  |  | . | 2433 | 1.3 | 0.560 | 1338 | 3383 | 1862 |
|  |  |  |  |  | . | 2409 | 1.4 | 0.603 | 1416 | 3300 | 1785 |
|  |  |  |  |  | . | 2384 | 1.5 | 0.646 | 1492 | 3220 | 1711 |
|  |  |  |  |  | . | 2361 | 1.6 | 0.689 | 1564 | 3144 | 1641 |
|  |  |  |  |  |  | 2337 | 1.7 | 0.732 | 1634 | 3069 | 1575 |
|  |  |  |  |  | . | 2313 | 1.8 | 0.775 | 1702 | 2998 | 1512 |
|  |  |  |  |  | . | 2290 | 1.9 | 0.818 | 1767 | 2929 | 1451 |
|  |  |  |  |  | . | 2267 | 2.0 | 0.861 | 1830 | 2863 | 1394 |

Input units are millions and kg - output in kilotonnes
Table 6.5.4 Blue Whiting. Prediction with management option table. Basis for 2002: F2002 = F2001; Recruitment: GM 1981-2000 = 15000 millions
MFDP version 1a
Run: ISVPA based
Blue whiting combined stock, 2003 WG Time and date: 15:09 04/05/03
Fbar age range: 3-7
Basis for 2003: F2003
Basis for 2003: F2003 $=F(2000-2002)=0.31$; Recruitment: GM 1981-2001 $=15000$ millions

| 2003 |  |  |  |  | 2004 |  |  |  |  | 2005 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 8536 | 5509 | 1 | 0.3087 | 1619 | 7511 | 5388 | 0.0 | 0.000 | 0 | 8228 | 6091 |
|  |  |  |  |  | . | 5349 | 0.1 | 0.031 | 175 | 8042 | 5882 |
|  |  |  |  |  | . | 5309 | 0.2 | 0.062 | 344 | 7862 | 5681 |
|  |  |  |  |  |  | 5270 | 0.3 | 0.093 | 509 | 7686 | 5488 |
|  |  |  |  |  | . | 5231 | 0.4 | 0.124 | 668 | 7516 | 5303 |
|  |  |  |  |  | . | 5193 | 0.5 | 0.154 | 823 | 7351 | 5124 |
|  |  |  |  |  | . | 5155 | 0.6 | 0.185 | 974 | 7191 | 4953 |
|  |  |  |  |  | . | 5117 | 0.7 | 0.216 | 1120 | 7036 | 4788 |
|  |  |  |  |  | . | 5080 | 0.8 | 0.247 | 1262 | 6885 | 4630 |
|  |  |  |  |  | . | 5043 | 0.9 | 0.278 | 1400 | 6739 | 4478 |
|  |  |  |  |  | . | 5006 | 1.0 | 0.309 | 1534 | 6597 | 4331 |
|  |  |  |  |  | . | 4969 | 1.1 | 0.340 | 1665 | 6459 | 4191 |
|  |  |  |  |  | . | 4933 | 1.2 | 0.370 | 1791 | 6325 | 4055 |
|  |  |  |  |  | . | 4897 | 1.3 | 0.401 | 1914 | 6194 | 3925 |
|  |  |  |  |  | . | 4861 | 1.4 | 0.432 | 2034 | 6068 | 3799 |
|  |  |  |  |  | . | 4826 | 1.5 | 0.463 | 2150 | 5945 | 3678 |
|  |  |  |  |  | . | 4791 | 1.6 | 0.494 | 2263 | 5826 | 3562 |
|  |  |  |  |  | . | 4756 | 1.7 | 0.525 | 2373 | 5710 | 3450 |
|  |  |  |  |  | . | 4722 | 1.8 | 0.556 | 2480 | 5597 | 3342 |
|  |  |  |  |  | . | 4687 | 1.9 | 0.587 | 2584 | 5487 | 3238 |
|  |  |  |  |  | . | 4653 | 2.0 | 0.617 | 2685 | 5381 | 3138 |

Input units are millions and kg - output in kilotonnes

Table 6.9.1 Blue whiting. Total landings, No. of samples, No. of fish measured and No. of fish aged by country and quarter for 2001.

| Country | Quarter | Landings (t) | No. Of samples | No. Fish measured | No. Fish aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 1 | 17311 | 7 | 344 |  |
|  | 2 | 6955 | 7 | 694 |  |
|  | 3 | 18950 | 34 | 1536 |  |
|  | 4 | 8063 | 5 | 202 |  |
|  | Total | 51279 | 53 | 2776 | 0 |
| Faroe Islands | 1 | 24016 | 5 | 1645 | 501 |
|  | 2 | 93274 | 12 | 3732 | 1200 |
|  | 3 | 80365 | 8 | 2278 | 800 |
|  | 4 | 7766 | 6 | 1773 | 600 |
|  | Total | 205420 | 31 | 9428 | 3101 |
| France | 1 | 4565 |  |  |  |
|  | 2 | 5307 |  |  |  |
|  | 3 | 4320 |  |  |  |
|  | 4 | 496 |  |  |  |
|  | Total | 14687 | 0 | 0 | 0 |
| Germany | 1 | 8080 |  |  |  |
|  | 2 | 8500 |  |  |  |
|  | 3 | 470 |  |  |  |
|  | 4 |  |  |  |  |
|  | Total | 17050 | 0 | 0 | 0 |
| Iceland | 1 | 4597 | 2 | 200 | 99 |
|  | 2 | 97543 | 29 | 2428 | 1188 |
|  | 3 | 159283 | 56 | 4546 | 2772 |
|  | 4 | 24116 | 16 | 1364 | 786 |
|  | Total | 285539 | 103 | 8538 | 4845 |
| Ireland | 1 | 6425 | 7 | 982 | 935 |
|  | 2 | 11396 |  |  |  |
|  | 3 |  |  |  |  |
|  | 4 | 4 |  |  |  |
|  | Total | 17825 | 7 | 982 | 935 |
| Norway | 1 | 253604 | 26 | 1300 | 1300 |
|  | 2 | 177532 | 25 | 1320 | 1320 |
|  | 3 | 107101 | 30 | 1750 | 500 |
|  | 4 | 33239 | 6 | 320 | 320 |
|  | Total | 571477 | 87 | 4690 | 3440 |
| Portugal | 1 | 274 | 71 | 7605 | 343 |
|  | 2 | 572 | 84 | 8090 | 333 |
|  | 3 | 586 | 72 | 7415 | 288 |
|  | 4 | 226 | 74 | 8185 | 413 |
|  | Total | 1659 | 301 | 31295 | 1377 |
| Russia | 1 | 32661 | 52 | 6465 | 1431 |
|  | 2 | 104456 | 50 | 6158 | 1312 |
|  | 3 | 83574 | 66 | 16076 | 592 |
|  | 4 | 69376 | 42 | 6603 | 194 |
|  | Total | 290067 | 210 | 35302 | 3529 |
| Scotland | 1 | 23606 | 1 | 295 | 0 |
|  | 2 | 2796 |  |  |  |
|  | 3 |  |  |  |  |
|  | 4 |  |  |  |  |
|  | Total | 26402 | 1 | 295 | 0 |
| Spain | 1 | 4713 | 56 | 4267 | 200 |
|  | 2 | 4827 | 54 | 4200 | 200 |
|  | 3 | 4525 | 44 | 2969 | 350 |
|  | 4 | 3442 | 57 | 5048 | 400 |
|  | Total | 17506 | 211 | 16484 | 1150 |
| Sweden | 1 | 1892 |  |  |  |
|  | 2 | 3467 |  |  |  |
|  | 3 | 7469 |  |  |  |
|  | 4 | 5722 |  |  |  |
|  | Total | 18550 | 0 | 0 | 0 |
| The Netherland | 1 | 15954 | 6 | 1647 | 150 |
|  | 2 | 18317 | 20 | 3070 | 500 |
|  | 3 | 3258 | 7 | 2388 | 175 |
|  | 4 |  |  |  |  |
|  | Total | 37529 | 33 | 7105 | 825 |
| Grand Total |  | 1554990 | 1037 | 116895 | 19202 |



Quarter 1


## Quarter 3



Quarter 2


Quarter 4

Figure 6.2.1. Total catches of blue whiting in 2002 by quarter and ICES rectangle. Grading of the symbols: small dots $10-100 \mathrm{t}$, white squares $100-1000 \mathrm{t}$, grey squares $1000-10000 \mathrm{t}$, and black squares $>10000 \mathrm{t}$. Excluding France, Sweden and Portugal.


Figure 6.2.2
Total catches of blue whiting in 2002 by ICES rectangle. Grading of the symbols: small dots 10 100 t , white squares $100-1000 \mathrm{t}$, grey squares $1000-10000 \mathrm{t}$, and black squares $>10000 \mathrm{t}$. Excluding France, Sweden and Portugal.


Figure 6.4.1.1.1 Distribution of blue whiting biomass (in 1000 tonnes) as estimated by R.V. "Johan Hjort" in April 2003.


Figure 6.4.1.1.2 Length and age distribution in the total (A) and spawning (B) stock of blue whiting in the area to the west of the British Isles as estimated by R.V. "Johan Hjort" in April 2003. Numbers in millions.

## Portuguese bottom trawl survey (Summer)



Spanish Bottom Trawl Surveys


Figure 6.4.2.1 Mean catch rates in the bottom trawl surveys from the southern area.

## CPUE Spanish pair trawlers



Figure 6.4.3.1 Blue whiting CPUE from Spanish Pair trawlers in ICES Div VIIIc and IXa (North)

## CPUE Norway



Figure 6.4.3.2. Blue whiting. Overall aggregated CPUE from the Norwegian directed fisheries in 1982-2002 (tonnes/hour).
















Abbreviations used in the figure legends: nsea $=$ norsea $=$ Norwegian Sea Acoustic survey; nspac $=$ norspa $=$ norspac $=$ Norway Spawning Acoustic survey; spcpue $=\mathrm{sp}=$ Spanish Pair Trawl CPUE; no $2002=$ without 2002 survey data; ruspac $=$ Russian Spawning Acoustic survey; hyb $=$ Laurec-Shephard hybrid model. The ISVPA $F_{3-7}$ values for the basic run and for the catch only run are derived from the selection pattern.

Figure 6.4.4.1 Comparison of assessments of the blue whiting stock by different methods (courtesy of the WG Methods, ICES 2003/D:3).


Figure 6.4.4.1.1. Blue whiting. Mean log catch by cohort and age, averaged for various periods.


Figure 6.4.4.1.2. Blue whiting. Log indices by year class and age group.


Figure 6.4.4.2.1. Blue Whiting. Results of data exploiration using AMCI software.


Figure 6.4.4.2.2. Blue Whiting: Residuals AMCI assessment WG 2003 Single fleets in combination with catch data (others removed)


Figure 6.4.4.2.3. Blue Whiting. Results of data exploiration using AMCI software.


Fig. 6.4.4.3.1
Blue whiting. ISVPA. Profiles of components of the loss function by sorces of information


Figure 6.4.4.3.2
Blue whiting. ISVPA.
Profiles of ISVPA loss function when signals from catch-at-age and all age-structurec indexes are taken with equal weights (1)
and when signal from catch-at-age from Norwegian SSB estimates are teken with equal weights (2)


Figure 6.4.4.3.3
Blue whiting. ISVPA
SSB esimates for ISVPA tuning on age-structured surveys or on SSB index


Figure 6.4.4.4.1. SSQ surface for the ICA assessment of Blue Whiting 2003.


Figure 6.4.4.4.2. Catch residuals and selection pattern in the ICA assessment of Blue Whiting 2003

| Lヨח二ings | Fishing Martalitu |
| :---: | :---: |
| Recruitment | stack size |

Figure 6.4.4.4.3. Standard plots for the ICA assessment of Blue Whiting 2003.





Figure 6.4.5.1.1. AMCI Blue whiting assessment 2003. Standard plots


Figure 6.4.5.1.2. Blue whiting: Residuals AMCI assessment WG 2003
Residuals of fleets in combined assessment


Figure 6.4.5.1.3. AMCI assessment of blue whiting 2003. SSQ for a range of terminal Fs


Figure 6.4.5.1.4. Final AMCI assessment on blue whiting. Fishing mortality at age and year.

Fishing mortality (5\%, 25\%, 50\%, 75\%, 95\%)


Recruitment, age 1 (5\%, 25\%, 50\%, 75\%, 95\%)


SSB (5\%, 25\%, 50\%, 75\%, 95\%)


Figure 6.4.5.1.5. An evaluation of uncertainty in AMCI assessment obtained by bootstrapping the residuals of the final AMCI assessment (1000 replicates).




Figure 6.4.5.1.6. Retrospective AMCI runs with terminal year set to 2000-2003.


Figure 6.4.5.2.1
Blue whiting.
Comparison of reported and theoretical catches (ISVPA)


[^6]

Figure 6.4.5.2.3
Blue whiting. ISVPA- derived estimates of $\mathrm{F}(\mathrm{a}, \mathrm{y})$




Figure 6.4.5.2.4
Blue whiting, ISVPA, retrospective runs, same options as in 2003




Figure 6.4.5.2.5
Blue whiting. ISVPA. Boostrap


Figure 6.4.5.3.1. Comparisons between final AMCI and ISVPA blue whiting assessments.



## $7.1 \quad$ The fishery

The catches of Icelandic summer-spawning herring from 1982-2002 are given in Tables 7.1.1 and 7.1.2. Discards were estimated at less than 1000 tonnes for the 2002/2003 season. The fishery started in September off east Iceland and will terminate at the end of April. The catch in September to mid-April was 88000 tonnes, but additional catch of a few hundred tonnes may be expected until the end of the season. The catch was taken with traditional purse-seines and pelagic trawls. The purse-seine fishery took place off both the east and west coast of Iceland in September through November and only minor purse-seine catches were taken later in the season. The pelagic trawl fishery started in October and the main fishery took place to the west of Iceland throughout the season. Because small herring was frequently present in the catches, especially at SE- and SW-Iceland, area closure was common during the 2002/2003 season.

A total of about 62000 tonnes were fished in the west of Iceland and about 25000 tonnes in the east. In the 1997/98 season $59 \%$ of the catch was taken by purse-seines, $78 \%$ in 1998/99, $61 \%$ in 1999/2000, $72 \%$ in 2000/2001 and only $38 \%$ in 2001/2002. In the $2002 / 2003$ season the trend of decreasing purse-seine catches continued with only $35 \%$ of the catches taken by purse-seine, the remainder by pelagic trawl.

The proportion used for reduction to meal and oil was $29 \%$ in 1997/98 and increased to $72 \%$ in 1998/99. This decreased again to $69 \%$ in $1999 / 2000$, to $64 \%$ in 2000/2001. Only $12 \%$ of the catch taken in the $2001 / 2002$ season was reduced to meal and oil, but this increased again in the 2002/2003 season to about $39 \%$. The remainder was either salted or frozen for human consumption.

Until 1990, the herring fishery took place during the last three months of the calendar year, but from 1990-2002 the autumn fishery has continued in January and early February of the following year. In 2003 the season was further extended to the end of April and in the summer of 2002 an experimental fishery for spawning herring with a catch of 5000 t was conducted at the south coast. In 1994 the fishery started in September. Therefore, all references to the years 1990-1993 imply seasons starting in October of that year, but after that in September. Landings, catches and recommended TACs since 1984 are given in thousand tonnes in Table 7.1.1.

### 7.2 Catch in numbers, weight-at-age and maturity

The catches of the Icelandic summer spawners in numbers-at-age for the period 1982-2002 are given in Table 7.1.2. Like last year age is now given as real age instead of rings, as in earlier WG reports.

During the 1995/96-1997/98 seasons, catches were mainly distributed on the 4 year classes from 1988-1991. On the other hand, catches during the 1998/99 and 1999/2000 seasons were dominated by the strong 1994 year class. In 2000/2001 the very strong 1994 and 1996 year classes were most abundant in the catch, but in 2001/2002 the 1996 year class was the most abundant. In 2002/2003 the 1998 and 1999 year classes were the most common in the catch followed by the 1996 and 1994 year classes.

The weight-at-age for each year is given in Table 7.2.1. As in previous years the weight-at-age in the next year (2003) is derived by a regression, whereas the mean weight in the previous year has been used to predict the weight-at-age for ages 3-9, using as input the weight-at-ages $2-8$ in the year before. Data for the regression included the period 19922002 as starting years. For one-year-old herring and 10+, a simple average of mean weights-at-age for the period 19982002 was used for the prediction. Weights-at-age for ages $2-8$ in the catch were obtained using the relationship:
$\mathrm{W}_{\mathrm{y}+1}-\mathrm{W}_{\mathrm{y}}=-0.2284 * \mathrm{~W}_{\mathrm{y}}+94.7855(\mathrm{~g})$
where $W_{y}$ and $W_{y+1}$ are the mean weight of the same year class in year $y$ and $y+1$, respectively.
The proportion mature-at-age is given in Table 7.2.2.

### 7.3 Acoustic surveys

The Icelandic summer-spawning herring stock has been monitored by annual acoustic surveys since 1973, Table 7.3.1. These surveys have been conducted in October-December or January. The 2002 survey started on the 24 November at the east coast of Iceland. On the traditional fishing grounds off the east coast the survey recorded only about 80000 t , while 180000 t were located in this region in 2001. West off Iceland very little adult herring was recorded during this
cruise, but the immature 1999 year classes were widely distributed. The estimate of the adult component was much less than expected and consequently a new survey, solely directed towards adult herring, was conducted to the west of Iceland in the first half of February 2003. During this survey a total of 385000 tonnes of adult herring were recorded so the sum of the estimated size of the adult stock was about 460000 t , which is somewhat less than expected (Gudmundsdottir, WD).

The 2002/2003 acoustic assessment surveys confirmed that the 1999 year class is large (Table 7.3.1).
The sum of results obtained in winter 2002/2003 acoustic surveys have been used as the basis for the present assessment of age 5 (age 6 on 1 January) and older herring (Table 7.3.1).

Jakobsson et al. (1993) formally tested whether it was feasible to maintain a one-to-one relationship between acoustic and VPA estimates of stock size. It was found that a modification of the target strength, from TS=21.7 $\log (\mathrm{L})-75.5 \mathrm{~dB}$ to $\mathrm{TS}=20 \log (\mathrm{~L})-72 \mathrm{~dB}$, gave a much better fit between the two data sets. The resulting target strength $\mathrm{TS}=20 \log (\mathrm{~L})-$ 72 dB was used to recalculate historic acoustic stock assessments. The latter target strength has been the basis of calculations of stock abundance from acoustic survey data since 1993.

### 7.4 Stock assessment

### 7.4.1 ADAPT-type of VPA

Using the results from the acoustic survey and the catch in numbers, a first estimate of $F$ was made. In this analysis, herring at age 6 (on 1 Jan 2003) and older have been grouped for estimating the fishing mortality for the oldest herring. For F on the oldest age group, an average F for ages $7-14$ was used. The resulting ADAPT-type run gave an F of 0.16 . The relationship between the VPA estimates and the acoustic estimates is shown in Figure 7.4.1.1.

A retrospective plot (Figure 7.4.1.2) shows that the terminal F values have been underestimated in the last 5 years. Therefore, like last year, the terminal F this year was increased by $19 \%$, which is the mean underestimate in the last 5 years, resulting in an F of 0.19 .

According to this assessment with the raised F of 0.19 , the spawning stock biomass was about 470000 t on 1 July 2002, which is about 80000 t lower than the estimate from last year.

### 7.4.2 AMCI assessment

The assessment program AMCI22 (Section 1.3.2) was also used. The objective function was a sum of the following partial objective function:

Log sum of squares of catches-at-age, weight 1
Log sum of squares of yearly yields, weight 1

Log sum of squares for the acoustic survey indices-at-age, weight 1
Different runs of AMCI were made with the following assumptions: As the icelandic summer-spawning herring is managed at a level close to 0.22 , then the fishing mortality in 2003 was assumed to be 0.22 and the recruitment 650 millions, which is close to the long-term geometric mean of 2-year-old recruiting herring. The yearly fishing mortality was split on quarters, assuming 0.05 in the third quarter of the year and 0.95 in the fourth quarter. Natural mortality of 0.1 was assumed.

In the first run, which was an imitation of the ADAPT-type of VPA and should be regarded as an exploratory run the selectivity was estimated for the first year, and in the last year the selectivity was fixed at 1 for ages 5 and older. The gainfactor was set at 1 (VPA type of run) for all ages and years between the first and the last year. In the survey fleet ages 5 to 14 were used at the survey time (age 6 to 15 at 1st of January the year after) with selection equal to 1 for all ages. By looking at the residuals in the catch and the observed vs. the modelled catch it is clear that there are two periods in the model. It fits quite well after 1992, but before the model assumes lesser catch. The ADAPT-type of VPA gave a terminal F of 0.16 and the AMCI model gives the terminal F of 0.23 , which is more in line with what is seen from the catch curves.

As AMCI is a forward-running model it seemed a natural next step to start running it from the first year where the selectivity was estimated and the selectivity pattern was allowed to change gradually for ages and years. The gainfactor was set to 0.5 for ages 2 and 3 and 0.2 for ages 4 and older. The selection in the survey fleet was estimated the first year, but kept constant after that. The age groups 1 to 14 were used in the survey fleet. By looking at the residuals big blocks both positive and negative were seen. These fitted with the time when big year classes were entering the fishery. Therefore several runs were made with different separable periods. The best run was obtained with four periods starting at 1981, 1983, 1986 and 1991. In Figure 7.4.2.1 a residual plot is shown both for the catches and the survey. A plot of the observed and modelled yield is also shown. A survey biomass index, both observed and modelled, was computed by using age groups 4 to 11 in the survey as they are in the fishable stock. The model fits the trends in the survey biomass estimate, except for the last year where the influence of the very strong 1999 year class is dominating. A retrospective run was also made and is shown in Figure 7.4.2.2. By comparing the retrospective plot from the ADAPT-type of VPA (Figure 7.4.1.2) and the retrospective plot from AMCI it can be seen that the AMCI model is more consistent for the last years. The working group decided to choose the AMCI assessment for the Icelandic summer-spawning herring this year. Fishing mortalities, stock numbers and a summary table are shown in Tables 7.4.1.1, 7.4.1.2 and 7.4.1.3.

According to this assessment the spawning stock biomass was about 475000 t at spawning time 2002, which is the same as derived by the ADAPT-type of VPA. The annual unweighted fishing mortality, F 5-15, amounted to 0.245 , which corresponds to a weighted F of 0.22 .

### 7.4.3 Comments to the assessment

Last year the assessment program ISVPA (Section 1.3.6) was also run. Due to time constraint it was not done this year.
At the last minute it was discovered that the residuals derived from the AMCI models showed a cyclic pattern, which could possibly mean some time trend in the model. It was decided to accept the assessment and to look closer into this pattern at a later time.

The AMCI model allows year-class multipliers to model targeted specific year classes. This feature should be explored further as the fishery of the Icelandic summer-spawning herring has been targeting the big year classes. This could not be done at the WG meeting because of lack of time.

### 7.5 Management consideration

During the last 20 years the Icelandic summer-spawning herring stock has been managed at levels corresponding fairly closely to fishing at $\mathbf{F}_{0.1}$. To derive the expected yield in the next years AMCI was run until 2004. The same selection was assumed as last year. According to the AMCI assessment, a catch of 108000 t in 2003 would exploit the stock at a fishing mortality level of a weighted F of 0.22 . The spawning stock biomass in 2003 is expected to be about 525000 t .

### 7.6 Comments on the PA reference points

The Working Group points out that managing this stock at an exploitation rate at or near $\mathbf{F}_{0.1}$ has been successful in the past. Thus the Working Group agreed in 1998 with the SGPAFM on using $\mathbf{F}_{\mathrm{pa}}=\mathbf{F}_{0.1}=0.22, \mathbf{B}_{\mathrm{pa}}=\mathbf{B}_{\lim } * \mathrm{e}^{1.645 \sigma}=300000 \mathrm{t}$ where $\mathbf{B}_{\mathrm{lim}}=200000 \mathrm{t}$. The Study Group on Precautionary Reference Points for Advice on Fishery Management met in February this year and concluded that it was not considered relevant to change the $\mathbf{B}_{\text {lim }}$ from 200000 t . The present working group agrees with this conclusion.

Table 7.1.1 Icelandic summer spawners. Landings, catches and recommended TACs in thousand tonnes.

| Year | Landings | Catches | Recommended <br> TACs |
| :--- | ---: | ---: | ---: |
| 1984 | 50.3 | 50.3 | 50.0 |
| 1985 | 49.1 | 49.1 | 50.0 |
| 1986 | 65.5 | 65.5 | 65.0 |
| 1987 | 73.0 | 73.0 | 70.0 |
| 1988 | 92.8 | 92.8 | 100.0 |
| 1989 | 97.3 | 101.0 | 90.0 |
| $1990 / 1991$ | 101.6 | 105.1 | 90.0 |
| $1991 / 1992$ | 98.5 | 109.5 | 79.0 |
| $1992 / 1993$ | 106.7 | 108.5 | 86.0 |
| $1993 / 1994$ | 101.5 | 102.7 | 90.0 |
| $1994 / 1995$ | 132.0 | 134.0 | 120.0 |
| $1995 / 1996$ | 125.0 | 125.9 | 110.0 |
| $1996 / 1997$ | 95.9 | 95.9 | 100.0 |
| $1997 / 1998$ | 64.7 | 64.7 | 100.0 |
| $1998 / 1999$ | 87.0 | 87.0 | 90.0 |
| $1999 / 2000$ | 92.9 | 92.9 | 100.0 |
| $2000 / 2001$ | 100.3 | 100.3 | 110.0 |
| $2001 / 2002$ | 95.3 | 95.3 | 125.0 |
| $2002 / 2003^{*}$ | 92.7 | 93.6 | 105.0 |

*Preliminary

Table 7.1.2 Icelandic summer spawners. Catch in numbers (millions) and total catch in weight (thous. tonnes). 1982 refers to season 1982/1983.

| Age/Year | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.454 | 1.475 | 0.421 | 0.112 | 0.100 | 0.029 | 0.879 |
| 3 | 19.187 | 22.499 | 18.015 | 12.872 | 8.172 | 3.144 | 4.757 |
| 4 | 28.109 | 151.718 | 32.244 | 24.659 | 33.938 | 44.590 | 41.331 |
| 5 | 38.280 | 30.285 | 141.354 | 21.656 | 23.452 | 60.285 | 99.366 |
| 6 | 16.623 | 21.599 | 17.043 | 85.210 | 20.681 | 20.622 | 69.331 |
| 7 | 38.308 | 8.667 | 7.113 | 11.903 | 77.629 | 19.751 | 22.955 |
| 8 | 43.770 | 14.065 | 3.916 | 5.740 | 18.252 | 46.240 | 20.131 |
| 9 | 6.813 | 13.713 | 4.113 | 2.336 | 10.986 | 15.232 | 32.201 |
| 10 | 6.633 | 3.728 | 4.517 | 4.363 | 8.594 | 13.963 | 12.349 |
| 11 | 10.457 | 2.381 | 1.828 | 4.053 | 9.675 | 10.179 | 10.250 |
| 12 | 2.354 | 3.436 | 0.202 | 2.773 | 7.183 | 13.216 | 7.378 |
| 13 | 0.594 | 0.554 | 0.255 | 0.975 | 3.682 | 6.224 | 7.284 |
| 14 | 0.075 | 0.100 | 0.260 | 0.480 | 2.918 | 4.723 | 4.807 |
| 15 | 0.211 | 0.003 | 0.003 | 0.581 | 1.788 | 2.280 | 1.957 |
| Catch | 56.528 | 58.867 | 50.304 | 49.368 | 65.500 | 75.439 | 92.828 |


| Age/year | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 3.974 | 11.009 | 35.869 | 12.006 | 0.869 | 6.225 | 7.411 |
| 3 | 22.628 | 14.345 | 92.758 | 79.782 | 35.560 | 110.079 | 26.221 |
| 4 | 26.649 | 57.024 | 51.047 | 131.543 | 170.106 | 99.377 | 159.170 |
| 5 | 77.824 | 34.347 | 87.606 | 43.787 | 87.363 | 150.310 | 86.940 |
| 6 | 188.654 | 77.819 | 33.436 | 56.083 | 25.146 | 90.824 | 105.542 |
| 7 | 43.114 | 152.236 | 54.840 | 41.932 | 28.802 | 23.926 | 74.326 |
| 8 | 8.116 | 32.265 | 109.418 | 36.224 | 18.306 | 20.809 | 20.076 |
| 9 | 5.897 | 8.713 | 9.251 | 44.765 | 24.268 | 19.164 | 13.797 |
| 10 | 7.292 | 4.432 | 3.796 | 9.244 | 14.318 | 17.973 | 8.873 |
| 11 | 4.780 | 4.287 | 2.634 | 2.259 | 3.639 | 16.222 | 9.140 |
| 12 | 3.449 | 2.517 | 1.826 | 0.582 | 0.878 | 2.955 | 7.079 |
| 13 | 1.410 | 1.226 | 0.516 | 0.305 | 0.300 | 1.433 | 2.376 |
| 14 | 0.844 | 1.019 | 0.262 | 0.203 | 0.200 | 0.345 | 0.927 |
| 15 | 0.348 | 0.610 | 0.298 | 0.102 | 0.100 | 0.345 | 0.124 |
| Catch | 101.000 | 105.097 | 109.489 | 108.504 | 102.741 | 134.003 | 125.851 |


| Age/Year | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 1.100 | 9.323 | 16.161 | 0.629 | 7.958 | 10.206 | 14.149 |
| 3 | 18.723 | 27.072 | 37.787 | 43.537 | 52.921 | 23.944 | 70.982 |
| 4 | 45.304 | 28.397 | 151.853 | 65.871 | 131.153 | 76.666 | 78.395 |
| 5 | 92.948 | 29.451 | 42.833 | 145.127 | 44.334 | 107.849 | 43.905 |
| 6 | 69.878 | 42.267 | 19.872 | 24.653 | 102.925 | 46.646 | 57.266 |
| 7 | 86.261 | 35.285 | 30.280 | 20.614 | 10.962 | 51.585 | 21.433 |
| 8 | 37.447 | 28.506 | 22.572 | 25.853 | 9.312 | 18.504 | 42.272 |
| 9 | 13.207 | 21.828 | 32.779 | 21.163 | 17.218 | 11.356 | 9.668 |
| 10 | 6.854 | 8.160 | 14.366 | 14.436 | 9.471 | 7.933 | 4.632 |
| 11 | 4.012 | 3.815 | 4.802 | 6.973 | 7.610 | 8.547 | 6.429 |
| 12 | 1.672 | 1.696 | 2.199 | 2.164 | 1.930 | 5.090 | 7.839 |
| 13 | 4.179 | 6.570 | 1.084 | 2.426 | 5.199 | 4.346 | 9.738 |
| 14 | 1.672 | 1.378 | 5.081 | 0.473 | 0.552 | 1.611 | 4.478 |
| 15 | 0.100 | 1.802 | 3.036 | 0.961 | 0.166 | 0.864 | 4.537 |
| Catch | 95.882 | 64.682 | 86.998 | 92.896 | 100.332 | 95.278 | 93.601 |

Table 7.2.1 Icelandic summer spawners. Weight at age (g).

| Age/Year | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 59 | 49 | 53 | 60 | 60 | 75 | 63 |  |
| 3 | 132 | 131 | 146 | 140 | 168 | 157 | 130 |  |
| 4 | 180 | 189 | 219 | 200 | 200 | 221 | 206 |  |
| 5 | 218 | 217 | 266 | 252 | 240 | 239 | 246 |  |
| 6 | 260 | 245 | 285 | 282 | 278 | 271 | 261 |  |
| 7 | 309 | 277 | 315 | 298 | 304 | 298 | 290 |  |
| 8 | 329 | 315 | 335 | 320 | 325 | 319 | 331 |  |
| 9 | 356 | 322 | 365 | 334 | 339 | 334 | 338 |  |
| 10 | 370 | 351 | 388 | 373 | 356 | 354 | 352 |  |
| 11 | 407 | 334 | 400 | 380 | 378 | 352 | 369 |  |
| 12 | 437 | 362 | 453 | 394 | 400 | 371 | 389 |  |
| 13 | 459 | 446 | 469 | 408 | 404 | 390 | 380 |  |
| 14 | 430 | 417 | 433 | 405 | 424 | 408 | 434 |  |
| 15 | 472 | 392 | 447 | 439 | 430 | 437 | 409 |  |


| Age/Year | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 63 | 75 | 74 | 63 | 74 | 67 | 69 | 78 |
| 144 | 119 | 139 | 144 | 150 | 135 | 129 | 140 |
| 190 | 198 | 188 | 190 | 212 | 204 | 178 | 166 |
| 232 | 244 | 228 | 232 | 245 | 249 | 236 | 208 |
| 276 | 273 | 267 | 276 | 288 | 269 | 276 | 258 |
| 317 | 286 | 292 | 317 | 330 | 302 | 292 | 294 |
| 334 | 309 | 303 | 334 | 358 | 336 | 314 | 312 |
| 346 | 329 | 325 | 346 | 373 | 368 | 349 | 324 |
| 364 | 351 | 343 | 364 | 387 | 379 | 374 | 360 |
| 392 | 369 | 348 | 392 | 401 | 398 | 381 | 349 |
| 444 | 387 | 369 | 444 | 425 | 387 | 400 | 388 |
| 399 | 422 | 388 | 399 | 387 | 421 | 409 | 403 |
| 419 | 408 | 404 | 419 | 414 | 402 | 438 | 385 |
| 428 | 436 | 396 | 428 | 420 | 390 | 469 | 420 |


| Age/Year | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | $2003^{*}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 62 | 78 | 64 | 58 | 78 | 80 | 70 |
| 3 | 137 | 147 | 143 | 158 | 140 | 149 | 156 |
| 4 | 197 | 184 | 211 | 214 | 217 | 202 | 210 |
| 5 | 234 | 213 | 236 | 256 | 242 | 245 | 250 |
| 6 | 270 | 246 | 268 | 284 | 281 | 275 | 284 |
| 7 | 299 | 286 | 300 | 326 | 294 | 311 | 307 |
| 8 | 323 | 314 | 318 | 333 | 309 | 325 | 335 |
| 9 | 342 | 341 | 349 | 366 | 339 | 347 | 345 |
| 10 | 358 | 351 | 347 | 383 | 350 | 383 | 367 |
| 11 | 363 | 354 | 377 | 402 | 367 | 390 | 379 |
| 12 | 373 | 350 | 359 | 405 | 375 | 402 | 392 |
| 13 | 412 | 372 | 403 | 422 | 403 | 442 | 407 |
| 14 | 394 | 400 | 408 | 406 | 426 | 463 | 414 |
| 15 | 429 | 437 | 445 | 444 | 425 | 453 | 433 |

[^7]Table 7.2.2 Icelandic summer spawners. Proportion mature-at-age.

| Age/Year | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 | 0.000 | 0.010 | 0.000 | 0.030 | 0.010 | 0.045 | 0.060 |
| 4 | 0.640 | 0.820 | 0.900 | 0.890 | 0.870 | 0.900 | 0.930 |
| 5 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 6 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 7 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 8 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 9 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 10 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 11 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 12 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 13 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 14 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 15 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |


| Age/Year | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 | 0.000 | 0.013 | 0.020 | 0.049 | 0.054 | 0.157 | 0.049 |
| 4 | 0.780 | 0.720 | 0.930 | 0.999 | 1.000 | 0.982 | 0.990 |
| 5 | 1.000 | 1.000 | 1.000 | 1.000 | 0.992 | 0.998 | 1.000 |
| 6 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 7 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 8 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 9 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 10 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 11 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 12 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 13 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 14 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 15 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |


| Age/Year | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | $2003^{*}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 | 0.160 | 0.265 | 0.074 | 0.279 | 0.101 | 0.190 | 0.190 |
| 4 | 0.925 | 0.935 | 0.879 | 0.831 | 0.981 | 0.734 | 0.849 |
| 5 | 0.989 | 0.995 | 0.977 | 0.992 | 0.997 | 0.898 | 0.962 |
| 6 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 7 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 8 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 9 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 10 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 11 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 12 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 13 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 14 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 15 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |

[^8]Table 7.3.1 Acoustic estimates (in millions) of the Icelandic summer spawning herring, 1974-2003.
The surveys are conducted in October-December or January. The year given is the following year, i.e. if the survey is conducted in the season 1973/1974, then 1974 is given.

|  | Ages |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 3 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 6+ |
| 1974 | -1 | 154 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1975 | -1 | 5 | 137 | 19 | 21 | 2 | 2 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | 25 |
| 1976 | -1 | 136 | 20 | 133 | 17 | 10 | 3 | 3 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | 33 |
| 1977 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1978 | -1 | 212 | 424 | 46 | 19 | 139 | 18 | 18 | 10 | -1 | -1 | -1 | -1 | -1 | -1 | 204 |
| 1979 | -1 | 158 | 334 | 215 | 49 | 20 | 111 | 30 | 30 | 20 | -1 | -1 | -1 | -1 | -1 | 260 |
| 1980 | -1 | 19 | 177 | 360 | 253 | 51 | 41 | 93 | 10 | -1 | -1 | -1 | -1 | -1 | -1 | 448 |
| 1981 | 625 | 361 | 462 | 85 | 170 | 182 | 33 | 29 | 58 | 10 | -1 | -1 | -1 | -1 | -1 | 482 |
| 1982 | -1 | 17 | 75 | 159 | 42 | 123 | 162 | 24 | 8 | 46 | 10 | -1 | -1 | -1 | -1 | 415 |
| 1983 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1984 | -1 | 171 | 310 | 724 | 80 | 39 | 15 | 27 | 26 | 10 | 5 | 12 | -1 | -1 | -1 | 214 |
| 1985 | -1 | 28 | 67 | 56 | 360 | 65 | 32 | 16 | 17 | 18 | 9 | 7 | 4 | 5 | 5 | 538 |
| 1986 | 201 | 652 | 208 | 110 | 86 | 425 | 67 | 41 | 17 | 27 | 26 | 16 | 6 | 6 | 1 | 718 |
| 1987 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1988 | 406 | 126 | 352 | 836 | 287 | 53 | 37 | 76 | 25 | 21 | 14 | 17 | 8 | 6 | 3 | 547 |
| 1989 | 370 | 725 | 181 | 249 | 381 | 171 | 42 | 23 | 30 | 16 | 10 | 9 | 5 | 3 | 2 | 692 |
| 1990 | -1 | 178 | 593 | 177 | 302 | 538 | 185 | -1 | -1 | -1 | 18 | -1 | -1 | -1 | -1 | 1043 |
| 1991 | 710 | 805 | 227 | 304 | 137 | 176 | 387 | 40 | 10 | 2 | -1 | -1 | -1 | -1 | -1 | 752 |
| 1992 | 465 | 745 | 850 | 353 | 273 | 94 | 81 | 210 | 32 | 11 | -1 | 17 | -1 | -1 | -1 | 718 |
| 1993 | 1418 | 254 | 858 | 687 | 160 | 99 | 87 | 44 | 92 | 39 | -1 | -1 | -1 | -1 | -1 | 521 |
| 1994 | 183 | 234 | 533 | 860 | 443 | 55 | 69 | 43 | 86 | 55 | 2 | -1 | 6 | -1 | -1 | 753 |
| 1995 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1996 | 845 | 98 | 165 | 515 | 316 | 361 | 166 | 110 | 52 | 29 | 16 | 27 | 19 | 8 | 2 | 1105 |
| 1997 | 266 | 792 | 65 | 139 | 459 | 280 | 410 | 150 | 101 | 50 | 35 | 15 | 65 | 32 | -1 | 1597 |
| 1998 | 1629 | 237 | 716 | 100 | 116 | 240 | 161 | 130 | 97 | 35 | 15 | 11 | 43 | 8 | 15 | 870 |
| 1999 | -1 | -1 | 188 | 790 | 240 | 101 | 73 | 47 | 77 | 47 | 10 | 10 | -1 | 22 | -1 | 627 |
| 2000 | 1069 | 527 | 740 | 296 | 606 | 99 | 71 | 164 | 108 | 98 | 15 | 44 | 5 | 13 | 7 | 1230 |
| 2001 | 2832 | 101 | 561 | 1069 | 323 | 609 | 30 | 31 | 38 | 13 | 18 | 6 | 9 | 4 | 1 | 1082 |
| 2002 | 561 | 942 | 247 | 187 | 265 | 173 | 302 | 69 | 48 | 55 | 54 | 16 | 18 | 1 | -1 | 1001 |
| 2003 | -1 | 1642 | 1800 | 549 | 221 | 257 | 159 | 273 | 97 | 44 | 43 | 13 | 32 | 2 | -1 | 1141 |

Table 7.4.1.1 Icelandic summer spawners. Fishing mortality-at-age.
Run id 20030505160344.463
Total yearly fishing mortalities at age

|  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.0012 | 0.0018 | 0.0034 | 0.0017 | 0.0009 | 0.0002 | 0.0001 | 0.0008 |
| 3 | 0.0193 | 0.0294 | 0.0901 | 0.0725 | 0.0510 | 0.0074 | 0.0069 | 0.0112 |
| 4 | 0.0967 | 0.1470 | 0.2625 | 0.1868 | 0.1788 | 0.0660 | 0.0766 | 0.0927 |
| 5 | 0.1430 | 0.2341 | 0.3488 | 0.2868 | 0.2684 | 0.1267 | 0.1664 | 0.1951 |
| 6 | 0.1890 | 0.2739 | 0.3246 | 0.2205 | 0.2461 | 0.2103 | 0.2482 | 0.3142 |
| 7 | 0.2802 | 0.4137 | 0.1608 | 0.1134 | 0.1229 | 0.4196 | 0.4868 | 0.5568 |
| 8 | 0.1617 | 0.2848 | 0.1592 | 0.1085 | 0.1089 | 0.3761 | 0.4716 | 0.5974 |
| 9 | 0.0951 | 0.1456 | 0.1266 | 0.0865 | 0.0833 | 0.3455 | 0.4387 | 0.5637 |
| 10 | 0.1554 | 0.2256 | 0.1058 | 0.0746 | 0.0778 | 0.3129 | 0.4250 | 0.5675 |
| 11 | 0.0557 | 0.1072 | 0.0852 | 0.0604 | 0.0632 | 0.3262 | 0.4421 | 0.5984 |
| 12 | 0.0910 | 0.1330 | 0.0741 | 0.0437 | 0.0541 | 0.3021 | 0.4162 | 0.5771 |
| 13 | 0.0819 | 0.1194 | 0.0177 | 0.0114 | 0.0171 | 0.1942 | 0.2363 | 0.3379 |
| 14 | 0.0208 | 0.0305 | 0.0423 | 0.0312 | 0.0312 | 0.3381 | 0.4868 | 0.6037 |
| 15 | 0.1177 | 0.1726 | 0.0965 | 0.0551 | 0.0791 | 0.3268 | 0.4618 | 0.6667 |
|  |  |  |  |  |  |  |  |  |
| Fref | 0.1265 | 0.1946 | 0.1401 | 0.0993 | 0.1047 | 0.2981 | 0.3891 | 0.5071 |

Total yearly fishing mortalities at age

|  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.0060 | 0.0093 | 0.0295 | 0.0212 | 0.0086 | 0.0159 | 0.0208 | 0.0092 |
| 3 | 0.0399 | 0.0390 | 0.1175 | 0.0922 | 0.0637 | 0.1377 | 0.1190 | 0.0860 |
| 4 | 0.0849 | 0.1056 | 0.2244 | 0.2029 | 0.1688 | 0.2357 | 0.2674 | 0.2229 |
| 5 | 0.1696 | 0.1777 | 0.3389 | 0.2771 | 0.2080 | 0.2946 | 0.2889 | 0.2540 |
| 6 | 0.2917 | 0.2981 | 0.3004 | 0.2810 | 0.2012 | 0.2918 | 0.2926 | 0.2544 |
| 7 | 0.4244 | 0.4539 | 0.3087 | 0.3149 | 0.2412 | 0.3214 | 0.3342 | 0.2912 |
| 8 | 0.4166 | 0.4365 | 0.4031 | 0.3408 | 0.2666 | 0.3610 | 0.3467 | 0.2851 |
| 9 | 0.4317 | 0.4407 | 0.2985 | 0.2888 | 0.2341 | 0.3675 | 0.3522 | 0.2892 |
| 10 | 0.4097 | 0.4353 | 0.2707 | 0.2500 | 0.1852 | 0.2749 | 0.2903 | 0.2337 |
| 11 | 0.5024 | 0.4686 | 0.2952 | 0.2553 | 0.1859 | 0.2681 | 0.2607 | 0.2157 |
| 12 | 0.4818 | 0.5057 | 0.1589 | 0.1382 | 0.1086 | 0.1593 | 0.1608 | 0.1178 |
| 13 | 0.2722 | 0.2864 | 0.2032 | 0.1509 | 0.1081 | 0.1769 | 0.1786 | 0.1426 |
| 14 | 0.4182 | 0.4955 | 0.1943 | 0.1749 | 0.1189 | 0.1928 | 0.2671 | 0.2545 |
| 15 | 0.4439 | 0.4548 | 0.2666 | 0.2197 | 0.1720 | 0.2649 | 0.2502 | 0.1926 |
|  |  |  |  |  |  |  |  |  |
| Fref | 0.3875 | 0.4048 | 0.2762 | 0.2447 | 0.1845 | 0.2703 | 0.2748 | 0.2301 |

Total yearly fishing mortalities at age

|  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.0080 | 0.0123 | 0.0060 | 0.0135 | 0.0121 | 0.0134 |  |  |
| 3 | 0.0434 | 0.0611 | 0.0563 | 0.0981 | 0.0836 | 0.0694 |  |  |
| 4 | 0.1655 | 0.2024 | 0.1788 | 0.1752 | 0.1978 | 0.2119 |  |  |
| 5 | 0.1921 | 0.2446 | 0.2283 | 0.1975 | 0.2177 | 0.2196 |  |  |
| 6 | 0.1916 | 0.2252 | 0.2066 | 0.2087 | 0.2280 | 0.2281 |  |  |
| 7 | 0.2183 | 0.2532 | 0.2331 | 0.1956 | 0.2061 | 0.2056 |  |  |
| 8 | 0.2095 | 0.2437 | 0.2176 | 0.1907 | 0.2317 | 0.2352 |  |  |
| 9 | 0.2189 | 0.2652 | 0.2367 | 0.2147 | 0.2418 | 0.2485 |  |  |
| 10 | 0.1824 | 0.2198 | 0.1855 | 0.1649 | 0.1695 | 0.1726 |  |  |
| 11 | 0.1618 | 0.1958 | 0.1641 | 0.1401 | 0.1568 | 0.1591 |  |  |
| 12 | 0.0913 | 0.1117 | 0.1009 | 0.0815 | 0.0889 | 0.1044 |  |  |
| 13 | 0.1231 | 0.1406 | 0.1312 | 0.1777 | 0.1815 | 0.1910 |  |  |
| 14 | 0.1842 | 0.3039 | 0.2339 | 0.1865 | 0.2236 | 0.2530 |  |  |
| 15 | 0.3967 | 0.5778 | 0.4673 | 0.3463 | 0.3887 | 0.6817 |  |  |
|  |  |  |  |  |  |  |  |  |
| Fref | 0.1973 | 0.2529 | 0.2187 | 0.1913 | 0.2122 | 0.2454 |  |  |

Table 7.4.1.2 Icelandic summer spawners. Stock size.
Run id 20030505160344.463
Stocknumbers at age,
in area 1
Data by 1. Jan., except at youngest age which are at recruitment time

| 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 810369.3 | 287400.6 | 295933.8 | 511834.8 | 1266989.2 | 715406.2 | 346667.0 | 508785.6 |
| 266368.6 | 732404.1 | 259594.2 | 266857.2 | 462346.1 | 1145353.1 | 647220.8 | 313632.6 |
| 202021.4 | 236424.0 | 643514.4 | 214648.3 | 224578.3 | 397545.0 | 1028743.7 | 581600.8 |
| 125025.7 | 165947.5 | 184688.6 | 447836.3 | 161130.8 | 169936.3 | 336732.4 | 862227.8 |
| 193221.1 | 98058.3 | 118814.2 | 117907.9 | 304190.1 | 111477.1 | 135460.9 | 257972.0 |
| 200022.7 | 144719.3 | 67466.9 | 77708.1 | 85577.4 | 215189.3 | 81741.0 | 95626.1 |
| 72980.1 | 136767.2 | 86583.7 | 51981.1 | 62772.7 | 68477.7 | 127987.6 | 45456.4 |
| 56188.2 | 56177.3 | 93077.3 | 66811.1 | 42199.7 | 50936.9 | 42540.3 | 72262.7 |
| 83950.0 | 46228.6 | 43944.3 | 74207.3 | 55444.0 | 35131.5 | 32625.5 | 24823.5 |
| 30335.0 | 65029.3 | 33382.8 | 35769.3 | 62321.0 | 46414.1 | 23246.7 | 19299.7 |
| 8860.3 | 25961.2 | 52860.7 | 27740.3 | 30467.3 | 52938.3 | 30307.6 | 13518.9 |
| 8845.0 | 7319.6 | 20565.5 | 44413.7 | 24027.2 | 26115.2 | 35409.8 | 18087.1 |
| 9655.2 | 7374.2 | 5877.6 | 18281.9 | 39733.1 | 21372.7 | 19458.9 | 25296.1 |
| 1281.8 | 3898.0 | 3838.0 | 3354.9 | 7686.8 | 16781.3 | 10059.1 | 6731.5 |

Stocknumbers at age,
in area 1
Data by 1. Jan., except at youngest age which are at recruitment time

|  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 499134.8 | 1000886.5 | 1272302.9 | 805503.8 | 773099.3 | 343011.3 | 321626.5 | 1155551.2 |
| 3 | 459978.9 | 448951.4 | 897253.5 | 1117765.9 | 713569.0 | 693521.7 | 305472.0 | 285042.4 |
| 4 | 280636.9 | 399926.9 | 390674.2 | 721847.8 | 922276.4 | 605804.9 | 546817.6 | 245397.8 |
| 5 | 479643.7 | 233254.5 | 325616.6 | 282441.6 | 533195.3 | 704876.4 | 433035.9 | 378673.0 |
| 6 | 641908.2 | 366306.3 | 176699.6 | 209941.9 | 193719.7 | 391870.4 | 475051.8 | 293500.8 |
| 7 | 170484.5 | 433856.9 | 246016.6 | 118396.4 | 143423.6 | 143333.5 | 264846.7 | 320789.0 |
| 8 | 49583.0 | 100909.5 | 249331.3 | 163479.9 | 78188.8 | 101963.6 | 94042.8 | 171571.3 |
| 9 | 22631.0 | 29577.3 | 59009.9 | 150754.1 | 105206.9 | 54191.0 | 64301.5 | 60165.1 |
| 0 | 37210.0 | 13298.5 | 17223.4 | 39615.9 | 102190.5 | 75324.2 | 33953.9 | 40909.0 |
| 1 | 12734.0 | 22351.8 | 7786.3 | 11887.9 | 27916.6 | 76832.4 | 51777.2 | 22982.6 |
| 2 | 9599.2 | 6972.1 | 12657.6 | 5244.3 | 8332.7 | 20975.7 | 53173.7 | 36098.1 |
| 3 | 6868.7 | 5365.2 | 3804.6 | 9770.3 | 4132.8 | 6763.6 | 16184.6 | 40966.4 |
| 4 | 11673.6 | 4734.0 | 3645.6 | 2809.7 | 7602.5 | 3356.2 | 5127.5 | 12248.6 |
|  | 6359.5 | 4327.7 | 2071.4 | 1688.0 | 1366.3 | 2906.5 | 1838.6 | 1970.9 |

Stocknumbers at age,
in area 1
Data by 1. Jan., except at youngest age which are at recruitment time

|  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 575372.8 | 991408.3 | 555279.4 | 480069.8 | 1135111.0 | 650000.0 | 650000.0 |
| 3 | 1036016.5 | 516492.6 | 886110.1 | 499440.7 | 428544.8 | 1014756.0 | 580341.2 |
| 4 | 236674.7 | 897644.3 | 439651.0 | 757905.3 | 409695.5 | 356659.1 | 856643.2 |
| 5 | 177683.9 | 181487.5 | 663394.5 | 332692.1 | 575592.7 | 304179.8 | 261082.3 |
| 6 | 265780.2 | 132670.0 | 128580.1 | 477724.9 | 247077.4 | 418909.4 | 220961.8 |
| 7 | 205920.3 | 198552.9 | 95843.1 | 94623.7 | 350856.7 | 177983.9 | 301740.0 |
| 8 | 216932.1 | 149777.4 | 139476.4 | 68690.0 | 70407.8 | 258338.6 | 131113.3 |
| 9 | 116728.3 | 159182.8 | 106209.2 | 101518.9 | 51361.0 | 50530.8 | 184756.6 |
| 10 | 40766.6 | 84857.9 | 110476.6 | 75845.9 | 74108.4 | 36493.2 | 35663.1 |
| 11 | 29300.4 | 30737.1 | 61632.1 | 83038.7 | 58192.8 | 56601.7 | 27785.7 |
| 12 | 16760.2 | 22552.6 | 22865.8 | 47326.5 | 65312.1 | 45015.2 | 43684.2 |
| 13 | 29032.0 | 13841.5 | 18250.1 | 18704.1 | 39470.0 | 54072.6 | 36693.2 |
| 14 | 32140.6 | 23226.7 | 10881.8 | 14482.6 | 14169.2 | 29786.1 | 40418.2 |
| 15 | 4091.5 | 10847.2 | 8544.7 | 5138.1 | 5758.4 | 5604.3 | 9550.7 |

Table 7.4.1.3 Icelandic summer spawners. Summary table.

Run id 20030505160344.463
SUMMARY TABLE

| Year | Recruits <br> age 2 | SSB | F | Catch <br> SOP |
| :--- | ---: | ---: | ---: | ---: |
| 1981 | 810369 | 247370 | 0.1265 | 39461 |
| 1982 | 287400 | 252699 | 0.1946 | 56472 |
| 1983 | 295933 | 272744 | 0.1401 | 58694 |
| 1984 | 511834 | 272198 | 0.0993 | 50132 |
| 1985 | 1266989 | 299543 | 0.1047 | 49309 |
| 1986 | 715406 | 302255 | 0.2981 | 65361 |
| 1987 | 346666 | 404279 | 0.3891 | 75295 |
| 1988 | 508785 | 462069 | 0.5071 | 92711 |
| 1989 | 499134 | 414622 | 0.3875 | 100868 |
| 1990 | 1000886 | 375107 | 0.4048 | 104854 |
| 1991 | 1272302 | 333458 | 0.2762 | 109235 |
| 1992 | 805503 | 395523 | 0.2447 | 108275 |
| 1993 | 773099 | 521610 | 0.1845 | 102513 |
| 1994 | 343011 | 537876 | 0.2703 | 133753 |
| 1995 | 321626 | 490488 | 0.2748 | 125673 |
| 1996 | 1155551 | 392202 | 0.2301 | 95722 |
| 1997 | 575372 | 375075 | 0.1973 | 64261 |
| 1998 | 991408 | 434105 | 0.2529 | 86849 |
| 1999 | 555279 | 438873 | 0.2187 | 92735 |
| 2000 | 480069 | 525385 | 0.1913 | 100406 |
| 2001 | 1135111 | 498631 | 0.2122 | 95352 |
| 2002 | 650000 | 474513 | 0.2454 | 93673 |
| 2003 | 650000 | 526206 | 0.2454 | 0 |
| $2004 *$ | 650000 | 519056 | 0.2454 | 0 |



Figure 7.4.1.1 Icelandic summer spawners. Acoustics estimates vs VPA stock numbers (at the $1^{\text {st }}$ of January)


Figure 7.4.1.2 Retrospective plot for ADAPT-type of VPA.


Observed catch and modelled catch


Figure 7.4.2.1 Residual plot from AMCI


Figure 7.4.2.2 A retrospective plot from AMCI.

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## Annex I

```
DETAILS OF DATA FILLING-IN
```



```
    >> (162) Norway
    Filling-in for record : ( 17)
Using Only
    >> (169) Norway 1 VIa
    Filling-in for record : ( 18) Denmark 2 VIa
Mean Weighted by Sampled Catches of:
    >> ( 62) Faroe Islands 2 VIa
    >> (126) The Netherlands 2 VIa
    >> (170) Norway 2 VIa
    Filling-in for record : ( 21) Denmark
Using Only
    >> (133) The Netherlands 1 VIIb
    Filling-in for record : ( 28) Ireland
Using Only
    >> (162) Norway 2 IVa
    Filling-in for record : ( 29) Ireland 1 VIa
Using Only
    >> (169) Norway 1 VIa
    Filling-in for record : ( 30) Ireland 2 VIa
Mean Weighted by Sampled Catches of:
    >> ( 62) Faroe Islands 2 VIa
    >> (126) The Netherlands 2 VIa
    >> (170) Norway 2 VIa
    Filling-in for record : ( 33) Ireland
Using Only
    >> (133) The Netherlands 1 VIIb
    Filling-in for record : ( 34) Ireland 2 VIIb
Using Only
    >> (133) The Netherlands 1 VIIb
    Filling-in for record : ( 41) Ireland 1 VIIj
Using Only
    >> (181) Norway 1 VIIk
    Filling-in for record : ( 42) Ireland 2 VIIj
Using Only
    >> (181) Norway 1 VIIk
    Filling-in for record : ( 45) Ireland 1 VIIIe
Using Only
    >> (181) Norway 1 VIIk
    Filling-in for record : ( 74) Germany 2 IIa
Mean Weighted by Sampled Catches of:
    >> ( 50) Faroe Islands 2 IIa
    >> ( 94) Iceland 2 IIa
    >> (114) The Netherlands 2 IIa
    >> (158) Norway 2 IIa
    Filling-in for record : ( 75) Germany 3 IIa
Mean Weighted by Sampled Catches of:
    >> ( 51) Faroe Islands 3 IIa
    >> ( 95) Iceland 3 IIa
```

```
    >> (159) Norway 3 IIa
    Filling-in for record : ( 77) Germany
Using Only
    >> (169) Norway 1 VIa
    Filling-in for record : ( 78) Germany 2 VIa
Mean Weighted by Sampled Catches of:
    >> ( 62) Faroe Islands 2 VIa
    >> (126) The Netherlands 2 VIa
    >> (170) Norway 2 VIa
    Filling-in for record : ( 79) Germany
Using Only
    >> (172) Norway 4 VIa
    Filling-in for record : ( 81) Germany
Using Only
    >> (173) Norway 1 VIb
    Filling-in for record : ( 85) Germany 1 VIIc
Mean Weighted by Sampled Catches of:
    >> ( 37) Ireland 1 VIIc
    >> (137) The Netherlands 1 VIIc
    >> (177) Norway 1 VIIc
    Filling-in for record : ( 89) Germany
Using Only
    >> (181) Norway 1 VIIk
    Filling-in for record : (105) Iceland 1 VIab+VIIbc
Mean Weighted by Sampled Catches of:
    >> (173) Norway 1 VIb
    >> (177) Norway 1 VIIc
    Filling-in for record : (106) Iceland 2 VIab+VIIbc
Mean Weighted by Sampled Catches of:
    >> (174) Norway 2 VIb
    >> (178) Norway 2 VIIc
    Filling-in for record : (119
    The Netherlands
    2 IVa
    The Netherlands
    1 VIb
    The Netherlands 3 VIIj
    1 VIIk
    Norway
    1 VIIgk+XII
Using Only
    >> (181) Norway
    Filling-in for record : (189) Scotland
    1 VIa
```

```
    Filling-in for record : (190) Scotland
    2 ~ V I a
Mean Weighted by Sampled Catches of:
    >> ( 62) Faroe Islands 2 VIa
    >> (126) The Netherlands 2 VIa
    >> (170) Norway 2 VIa
    Filling-in for record : (193) Scotland
    1 VIIb
Using Only
    >> (133) The Netherlands
    1 VIIb
    Filling-in for record : (197) Scotland
Mean Weighted by Sampled Catches of:
    >> ( 37) Ireland 1 VIIc
    >> (137) The Netherlands 1 VIIc
    >> (177) Norway 1 VIIc
    Filling-in for record : (201) France
Mean Weighted by Sampled Catches of:
    >> ( 57) Faroe Islands 1 Vb
    >> (173) Norway 1 VIb
    >> (177) Norway 1 VIIc
    Filling-in for record : (202) France
Mean Weighted by Sampled Catches of:
    >> ( 58) Faroe Islands 2 Vb
    >> (174) Norway 2 VIb
    Filling-in for record : (203) France
Using Only
    >> ( 59) Faroe Islands 3 Vb
    Filling-in for record : (204) France
Using Only
    >> ( 60) Faroe Islands 4 Vb
    Filling-in for record : ( 2) DENMARK
Using Only
    >> ( 14) Norway 2 IIa
    Filling-in for record : ( 3) DENMARK
Using Only
    >> ( 15) Norway 3 IIa
    Filling-in for record : ( 4) DENMARK
Using Only
    >> ( 16) Norway
        4 IIa
    Filling-in for record : ( 5) DENMARK
Using Only
    >> ( 17) Norway
    1 IVa
    Filling-in for record : ( 7) DENMARK
Using Only
    >> ( 19) Norway
    3 IVa
    Filling-in for record : ( 8) DENMARK
Using Only
    >> ( 20) Norway
    Filling-in for record : ( 10) DENMARK
    2 IVb
Using Only
```

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    >> ( 18) Norway
    Filling-in for record : ( 11)
Using Only
    >> ( 19) Norway
    Filling-in for record : ( 12)
Using Only
    >> ( 20) Norway
    Filling-in for record : ( 21)
Using Only
    >> ( 17) Norway
    Filling-in for record : ( 22)
Using Only
    >> ( 18) Norway
    Filling-in for record : ( 23)
Using Only
    >> ( 19) Norway
    Filling-in for record : ( 24)
Using Only
    >> ( 20) Norway
    Filling-in for record : ( 27)
Using Only
    >> ( 19) Norway
    Filling-in for record : ( 29)
Using Only
    >> ( 13) Norway
    Filling-in for record : ( 30)
Using Only
    >> ( 14) Norway
    Filling-in for record : ( 31)
Using Only
    >> ( 15) Norway
    Filling-in for record : ( 32)
Using Only
    >> ( 16) Norway
    Filling-in for record : ( 33)
Using Only
    >> ( 13) Norway
    Filling-in for record : ( 34)
Using Only
    >> ( 14) Norway
    Filling-in for record : ( 35) Sweden
Using Only
    >> ( 15) Norway
    Filling-in for record : ( 36)
Using Only
```

```
    >> ( 16) Norway
    Filling-in for record : ( 37)
Using Only
    >> (17) Norway
    Filling-in for record : ( 38)
Using Only
    >> ( 18) Norway
    Filling-in for record : ( 39)
Using Only
    >> ( 19) Norway
    Filling-in for record : ( 40)
Using Only
    >> ( 20) Norway
    Filling-in for record : ( 41)
Using Only
    >> ( 17) Norway
    Filling-in for record : ( 42)
Using Only
    >> ( 18) Norway
    Filling-in for record : ( 43)
Using Only
    >> ( 19) Norway
    Filling-in for record : ( 44)
Using Only
    >> ( 20) Norway
    Filling-in for record : ( 5) PORTUGAL
Using Only
    >> ( 1) Spain 1 VIIIc+IXa
    Filling-in for record : ( 6) PORTUGAL
Using Only
    >> ( 2) Spain 2 VIIIc+IXa
    Filling-in for record : ( 7) PORTUGAL
Using Only
    >> ( 3) Spain
    Filling-in for record : ( 8)
Using Only
    >> ( 4) Spain
4 VIIIc+IXa
```


[^0]:    1 Includes also by-catches of adult herring in other fisheries
    2 In 1972, there was also a directed herring 0 -group fishery
    3 Includes 26,000 tof immature herring (1983 year class) fished by USSR in the Barents Sea
    4 Preliminary, as provided by Working Group members
    ${ }^{5}$ Details of catches by fishery and ICES area given in ICES 1999
    ${ }^{6}$ Details of catches by fishery and ICES area given in ICES 2000
    ${ }^{7}$ Details of catches by fishery and ICES area given in ICES 2001
    ${ }^{8}$ Details of catches by fishery and ICES area given in ICES 2002
    ${ }^{9}$ Details of catches by fishery and ICES area given in Tables 3.2.1.3-3.2.1.5

[^1]:    

[^2]:    ${ }^{1}$ Preliminary data

[^3]:    *Preliminary

[^4]:    ${ }^{1}$ Three of these were split into two periods.
    ${ }^{2}$ The youngest age group in the Spanish CPUE tuning fleet was down-weighted.

[^5]:    ${ }^{3}$ AMCI 2.3 was used for scanning over terminal F's.
    ${ }^{4}$ Age 1 was down-weighted by factor 0.01 .

[^6]:    Figure 6.4.5.2.2
    Blue whiting, ISVPA. Residuals for catch-at-age and surveys

[^7]:    * Predicted

[^8]:    * Predicted (mean of 2000-2002)

