## Report of the

# Study Group on Incorporation of Process Information into Stock Recruitment Models 

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### 1.2 Terms of reference

The Study Group on Incorporation of Process Information into Stock Recruitment Models [SGPRISM] (Chair: C. O'Brien, UK) met in Lowestoft, UK from 14-18 January 2002 to:
a) further consider the two case studies (North Sea cod and Bay of Biscay anchovy);
b) further develop the modelling and testing of process and recruitment relationships for incorporation into management procedures;
c) identify the sources of data necessary for the determination of the reproductive potential of fish stocks and specify a protocol for their collection within existing co-ordinated surveys.

SGPRISM will report for the attention of the Oceanography and Resource Management Committees. The report will be brought to the attention of the Working Group on Recruitment Processes.

### 1.3 Scientific justification for the Study Group

The relationship between spawning stock and recruitment is fundamental to the scientific approach to fisheries management. Considerations of environmental factors can make a difference to how one might manage a stock. Simulation models can play an important role in helping identify whether and where benefits to management are most likely to accrue and therefore where it would be best to focus attention in terms of other (e.g., process) studies. Results from simulation studies should be used to guide biological studies. Short-term focused studies aimed at identifying likely mechanisms are also crucial, but results from such studies can only be put to full use with information from longer-term observations. There is also benefit in long-term studies of the environment and underlying processes so that one is prepared if something unexpected happens. For example, if a process study reveals a strong relationship with some environmental variable, then it would be possible to incorporate this immediately if historic data are already available.

The Study Group has afforded an opportunity for biologists and stock assessment practitioners to meet under a common theme and to begin the much needed process of integrating biological knowledge and stock assessment methods/techniques.

The first meeting of the Study Group (ICES 2000a) had concentrated mainly on environmental issues as drivers of recruitment variability; whilst the second meeting jointly considered possible environmental and biological causes for recruitment fluctuations (ICES 2001a). Whilst the Study Group had addressed its terms of reference at both meetings, it was felt that much work still remained to be undertaken and that a third, and final, meeting should be held.

The terms of reference (ToR) are addressed within the four main sections of the report. Specifically, ToR a) is addressed within Sections $2-5$ of the report, ToR b) is addressed within Sections 3 and 5, and ToR c) is addressed in Section 2.

For a given level of spawning stock biomass there is often considerable variation in recruitment. This variation is frequently attributed to environment effects on survival. However, there is increasing evidence that the age, size and spatial structure of the spawning stock and the physiological condition of spawners can influence the number of surviving recruits (ICES 2001a). Section 2 of this current report reviews approaches for quantifying the reproductive potential of individuals and stocks; building on the work of a recent NAFO Scientific Council Working Group on Reproductive Potential (BD1).

In Section 3, a limited number of area-based case studies are presented that illustrate a sequential introduction of biological processes into estimates of stock biomass. Recruitment is viewed from both a biological and a stock assessment perspective. Potential impacts of spawning characteristics on reference points for fishery management are reviewed and discussed in the context of the Precautionary Approach (PA). Results from a modelling of the effects of environment variability on North Sea cod are presented in Section 4, together with the investigation of modeldiagnostics and a temperature-induced characterisation of the weight-at-age relationship. Data for North Sea plaice were also examined in parallel to investigate systematic changes in condition factor.

The further development of stock assessment software tools, as proposed by the ICES Working Group on Methods on Fish Stock Assessments [WGMG] (ICES 2002c), is discussed in Section 5 in the context of the WGMTERM software for medium-term projections. Time series characterisations of historical weight-at-age are presented for a selection of North Sea stocks, including North Sea cod. A model of spawning-stock structure in Northeast Arctic cod is presented to investigate those components of the spawning population, which have had the greatest influence on subsequent recruitment. The model was applied during this meeting of SGPRISM to North Sea cod and herring, in order to ascertain its general applicability and whether it would be beneficial to use it more widely.

Further work and the relevance of the Study Group to similar activities within ICES and NAFO are discussed in Section 6.

The low accuracy of the environmental indices as recruitment predictors makes it impossible at present to estimate the population abundance one-year in advance (ICES 2000a; ICES 2001a; ICES 2002b). Therefore any use of those indices is postponed so far for the provision of forecast advice to managers until better predictive power of the environmental stock-recruitment models is achieved. At a recent meeting of the SPACC/IOC Study Group on the Use of Environmental Indices in the Management of Pelagic Populations (Cape Town, South Africa, 3-5 September 2001), it was proposed that a simulation analysis be conducted to evaluate the benefits of using environmentally linked recruitment predictors in the management of anchovy stocks. This work will be conducted later in 2002 and the results will be pertinent to ICES.

## 2 REPRODUCTIVE POTENTIAL

### 2.1 Introduction

The stock-recruitment (S-R) relationship assumes that spawner biomass (SB) is directly proportional to the reproductive potential of the stock, i.e., that the relative fecundity (number of eggs produced per unit weight) of the stock is constant both temporally and spatially. This assumption has long been recognized as invalid (Oosthuizen and Daan 1974; Ware 1980), however, until recently there has been no concerted effort directed towards replacing SB with more sensitive measures of reproductive potential.

In the case of Atlantic cod processes such as skipped spawning and/or atresia (Kjesbu et al. 1991; Witthames and Greer Walker 1995; Ma et al. 1998; Marshall et al. 1998; Bromley et al. 2000; Rideout et al. 2000) contribute to the lack of proportionality between total egg production and SB over time. As both responses are observed in poor condition fish, a high degree of interannual variability in food abundance and condition are likely to increase the divergence between reproductive potential and SB. Reproductive potential is also affected by shifts in size composition because large/old spawners have higher relative fecundities than small/young spawners (Marshall et al. 1998; Marteinsdottir et al. 2000). The progressive loss of large spawners in stocks experiencing sustained high fishing mortalities (Trippel 1999) can therefore reduce the total egg production disproportionately to the reduction in spawner biomass.

As a result of increased knowledge of variability in relative fecundity of individuals and stocks, the S-R relationships for several stocks are being re-evaluated using alternative indices of reproductive potential. The approach taken to reestimating reproductive potential varies according to the type of data available. A limited number of stocks have fecundity data that is being used to hindcast potential total egg production by the stock (Köster et al. 2001d, Marshall et al. WD9). Bioenergetic approaches to quantifying the reproductive potential of stocks have also been developed (Painting et al. 1998; Henderson et al. 2000; Marshall et al. 1999, 2000). These take advantage of historical databases derived from industrial sources (e.g., liver condition, oil meal ratios). Long time series describing spawner condition have also been used as proxies for the reproductive potential of individual spawners (Marshall and Frank 1999; Blanchard 2000). Basic demographic data (e.g., numbers-at-age) have been used to construct age diversity indices that can potentially be used as simple proxies for reproductive potential (Lambert 1990; Marteinsdottir and Thorarinsson 1998; Secor 2000).

Stocks span an information gradient in the sense of having variable quantities of biological data that are relevant to estimating stock reproductive potential (Figure 2.1). In the worst case, reproductive potential is quantified by using time-invariant maturity- and weight-at-age values to estimate spawner biomass. These estimates are driven solely by variation in spawner quantity. Incorporating year-specific values for maturity- and weight-at-age increases the sensitivity of spawner biomass to the effect of interannual variation in growth and condition on reproductive potential. Incorporating year-specific fecundity information into estimates of reproductive potential (e.g., total egg production) can also increase the sensitivity of estimates (Armstrong et al. WD7). Ongoing research is being undertaken with the goal of incorporating interannual differences in egg quality, egg mortality (e.g., predation) or spatial/temporal origin of eggs.

To provide an overview of available information and data resources the NAFO Working Group on Reproductive Potential is compiling tables summarizing the available data and data sources for selected stocks within NAFO and ICES jurisdictions'. These tables will soon be available, both through the NAFO website and in the Journal of the Northwest Atlantic Fishery Science. The tables for Northeast Arctic cod (Appendices A-D) and North Sea cod (Appendices E-H) are included in this report to illustrate the range of data available for data-rich and data-poor stocks, respectively. The analytical assessment is providing basic data about stock size (total abundance) and age/size composition for both stocks. However, the relative scarcity of data pertaining to reproductive potential for the North Sea cod stock is immediately apparent from comparing Appendix A and Appendix E with respect to the number of years having data on sex ratio, maturity, weight, condition and fecundity.

In the case of data-poor stocks alternative indices of reproductive potential are necessarily restricted to proxies that could be derived directly from the analytical assessment (e.g., age diversity indices) or limited time series. The approach taken at this meeting of SGPRISM was to contrast Northeast Arctic cod and North Sea cod in terms of the available data resources, types of indices of reproductive potential that can be estimated, and how the indices can potentially be used in stock management. Given the lack of knowledge, SGPRISM undertook an exploratory exercise for the North Sea cod.

### 2.2 Northeast Arctic cod

Northeast Arctic cod is an example of a stock having several historical databases that are relevant to estimating reproductive potential (Appendices A-D). A simulation analysis suggested that the dynamic range of SB is less than the range in total egg production (Marshall et al. 1999). Consequently several alternative measures of reproductive potential, including total egg production (Marshall et al. WD9) and total lipid energy (Marshall et al. 2000), are being estimated for the assessment time period (1946-present) and being compared to both SB and recruitment.

### 2.2.1 Estimating reproductive potential

Until recently, values of SB for Northeast Arctic cod were estimated using knife-edge maturity ogives (1946-1981) and constant weight-at-age values (1946-1982). The resulting S-R relationship (Figure 2.2a) was therefore insensitive to the effect of growth variation on reproductive potential. In 2001, the ICES Arctic Fisheries Working Group [AFWG] compiled historical data from Russian and Norwegian sources and developed new time series for maturity- and weightat age (ICES 2001b). Replacing the constant values led to a substantial downward revision of SB such that the longterm mean decreased from $577,425 \mathrm{t}$ (1946-1999) to $372,934 \mathrm{t}$ (1946-2000; ICES 2001b). In spite of this major change in the SB time series, the major change to the S-R relationship was a shift of outlying post-war values towards lower SB values (Figure 2.2b).

Concurrent to the revision of the SB time series, preliminary estimates of total egg production were made (Marshall et al. WD9). These estimates were calculated using a general fecundity model developed from field observations made during a time period when the condition of Barents Sea cod decreased rapidly due to the collapse of the Barents Sea
capelin stock (Kjesbu et al. 1998). In the general model the year effect on the fecundity/length relationships was represented using condition. The relationship between total egg production and recruitment shows that the high recruitment values for 1963, 1964 and 1970 are more consistent with the level of egg production (Figure 2.2c) compared to levels of SB (Figure 2.2b). The SB and total egg production time series show major differences with total egg production being higher in the 1970's than in the 1990's while the opposite holds true for SB estimates (Figure 2.3).

Identifying the factors causing discrepancies between SB and total egg production is a high priority for future research. There are several refinements planned for estimates of total egg production that will hopefully increase their sensitivity to true variation in reproductive potential. These include:

- improvements to the year-specific age/length keys used to convert VPA numbers-at-age to numbers-at-length;
- improvements to the general fecundity model to better utilize environmental information, e.g., capelin stock biomass or liver condition index;
- checks on the consistency between time series of growth parameters (maturity, weight and fecundity) used in estimates of SB and total egg production; and
- corrections for size- or condition-dependent effects on egg quality.


### 2.2.2 Incorporating this information into stock management

The AFWG is currently investigating the appropriateness of the biomass reference points in light of the revised maturity- and weight-at-age times series (Fig 2.2b). However, it should also be recognized that reference points for SB will have a degree of uncertainty resulting from any divergence between SB and total egg production (Figure 2.3). Reference points that are explicit for reproductive potential (e.g., limit and threshold total egg production) will be developed further in future research. Classifications of stock status using these reference points will be compared to the classifications made using $\mathrm{B}_{\mathrm{pa}}$. Medium-term stock projections would need modification to take into account reference points for reproductive potential and the resulting stock projections would need to be compared to projections of SB. Inter-sessional work towards this goal is planned for the summer of 2002 through a collaboration between scientists from Aberdeen and Bergen. This work will be presented as a case study at the proposed meeting of the Study Group on Growth, Maturity and Condition Indices in Stock Projections [SGGROMAT] that is planned for December 2002 (see Section 6.1). This work will extend current stock projection methods with existing biological data and models for Northeast Arctic cod.

### 2.3 North Sea cod

The data available to construct more sensitive measures of reproductive potential for North Sea cod on time scales that are comparable to the assessment were initially thought to be limited (Appendices E-H), particularly when compared to a data-rich stock such as Northeast Arctic cod (Appendices A-D). Closer investigation revealed there to be limited quantities of relevant data. Cod maturity data from the IBTS for 1980-1995 are given in Cook et al. (1999). Maturity ogives separated by sex are reported in Rijnsdorp et al. (1991) for the 1985-1989 period. A historical comparison of the changes of length-maturity ogives is given in Oosthuizen and Daan (1974) extending from the 1890s to the 1970s. It shows a shift in $\mathrm{L}_{50}$ from approximately 75 cm to just above 50 cm in length at maturity over this long-term period (Figure 2.4). Condition data are available for 11 years from English groundfish surveys (see Section 4.3). Fecundity data exist for 1970-1972 (Oosthuizen and Daan 1974) and for 1987-1988 (Heessen, unpublished, reported in Rijnsdorp et al. 1991). Mean dry weights of eggs are available from eggs collected from ripe and running females in 1969, 1971, 1972 and 1987 (Rijnsdorp et al. 1991). Estimates of egg production have also been made according to egg surveys carried out in the southern Bight during 1970-1974 (Daan 1981).

### 2.3.1 Estimating reproductive potential

Using the available data this meeting of SGPRISM estimated three alternative indices of reproductive potential (age diversity, female-only biomass and condition) for North Sea cod.

### 2.3.1.1 Age diversity

Stocks experiencing sustained high fishing mortalities typically exhibit truncated age structure (Trippel 1999). Age diversity indices have been used as an index of reproductive potential for Icelandic cod (Marteinsdottir and Thorarinsson 1998) and striped bass in Chesapeake Bay (Secor 2000). As it is easily estimated from basic demographic data, the utility of this index was examined for North Sea cod. The numbers of mature fish at each age, obtained using
the proportion mature as used in the assessments (Table 2.1a), were taken from the 2001 XSA assessment (ICES 2002a). A Shannon diversity index (H) was then estimated to the numbers of mature fish as follows:

$$
\mathrm{H}=1 / n\left(\log (n)-\sum_{i=1}^{k} f_{i} \log \left(f_{i}\right)\right)
$$

Eq. 1
where $k$ is the number of age groups, $n$ is the total number of mature fish in all age groups and $f_{i}$ is the number of mature fish in each age group (Shannon 1948). The assessment for North Sea cod used fixed values of proportion mature to estimate spawner biomass. Values of H were estimated using fixed proportions mature and annual estimates of the proportion mature derived from annual surveys (Cook et al. 1999, Armstrong et al. WD 7, Table 2. 1b).

The age diversity of North Sea cod has shown a long-term decline since 1963 (Figure 2.5). Since 1993, age diversity has been below the long term mean value. Using values of H estimated with variable annual proportions mature suggests a steeper decline although the data series is shorter. In the 1990s fishing pressure was very high on the North Sea cod stocks and there has been an accompanying decline in the SSBs. Hence it is not surprising that the age diversity has declined in recent years.

Age diversity indices estimated using time invariant proportions mature show no relationship with recruitment (Figure 2.5). However, the use of year-specific proportions mature suggests that there is a relationship between age diversity and recruitment (Figure 2.6, $\mathrm{r}^{2}=0.31$ ). The relationship is stronger than that observed for Icelandic cod $\left(\mathrm{r}^{2}=0.15\right.$, Marteinsdottir and Thorarinsson 1998). A multiple linear regression of recruitment (R) on age diversity (H) and $\log _{\text {e }}$ SSB gives a weak positive relationship $\left(\mathrm{r}^{2}=0.44, \mathrm{p}=0.06\right)$ :
$R=25.06 \times 10^{6}-37.60 \times 10^{6} H-2.23 \times 10^{6} \log _{e} S S B+3.38 \times 10^{6}\left(H^{*} \log _{e} S S B\right)$
Eq. 2

In terms of assessing reproductive potential, it appears that age diversity is a useful tool for describing major changes in the population age structure that will impact on reproduction (Marteinsdottir and Thorarinsson 1998, Secor 2000). For North Sea cod the correlation between age diversity and recruitment was only significant when the variable proportion mature in the stock were used to estimate H . Time invariant maturity ogives give an incorrect perception of age diversity as well as other aspects of the stock dynamics.

Table 2.1.a. Proportion of fish mature time invariant used in the 2001 XSA stock assessments of North Sea cod.

## a)

| Age | Proportion <br> mature |
| :---: | :---: |
| 1 | 0.01 |
| 2 | 0.05 |
| 3 | 0.23 |
| 4 | 0.62 |
| 5 | 0.86 |
| 6 | 1.00 |
| 7 | 1.00 |
| 8 | 1.00 |
| 9 | 1.00 |
| 10 | 1.00 |
| $+\operatorname{grp}$ | 1.00 |

Table 2.1.b. Survey derived estimates of proportion mature of North Sea cod, both sexes combined (Cook et al. 1999).
b)

|  | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | + grp |  |  |  |  |  |
| 1980 | 0.01 | 0.1 | 0.08 | 0.56 | 1 | 0.88 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1981 | 0.01 | 0.04 | 0.16 | 0.52 | 0.85 | 0.87 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1982 | 0 | 0.03 | 0.17 | 0.64 | 0.93 | 1 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1983 | 0 | 0.03 | 0.2 | 0.5 | 0.96 | 0.99 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1984 | 0 | 0.06 | 0.25 | 0.42 | 0.81 | 0.98 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1985 | 0 | 0.04 | 0.1 | 0.45 | 0.56 | 0.91 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1986 | 0 | 0.09 | 0.18 | 0.35 | 0.63 | 0.9 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1987 | 0 | 0.02 | 0.15 | 0.6 | 0.76 | 0.98 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1988 | 0 | 0.05 | 0.27 | 0.53 | 0.9 | 0.95 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1989 | 0 | 0.13 | 0.24 | 0.59 | 0.77 | 0.97 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1990 | 0.01 | 0.19 | 0.54 | 0.73 | 0.93 | 0.99 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1991 | 0 | 0.09 | 0.39 | 0.5 | 0.89 | 0.98 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1992 | 0 | 0.21 | 0.52 | 0.85 | 0.85 | 1 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1993 | 0 | 0.07 | 0.42 | 0.7 | 0.88 | 0.96 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1994 | 0 | 0.14 | 0.43 | 0.77 | 0.98 | 1 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |
| 1995 | 0.01 | 0.05 | 0.3 | 0.82 | 1 | 0.97 | 1 | 1 | 1 | 1 | 1 |  |  |  |  |  |

### 2.3.1.2 Condition

Condition indices reflect the amount of available energy reserves of the fish (Kjesbu et al. 1991). Adult female condition affects the quantity (number of eggs produced by individuals) and possibly the quality and survival of offspring (Kjesbu et al. 1991, Marteinsdottir and Steinarsson, 1998). Fluctuations in temperature and food can contribute to inter-annual variability in condition. For example, in Northeast Arctic cod, condition is positively correlated to mean annual bottom temperature and, particularly, capelin stock biomass (Kjesbu et al. 1998, Yaragina and Marshall 2000).

Condition can be measured using a bioenergetic approach (e.g., lipid analyses, liver condition index) or more simply based on the relationship between weight and length (Lambert and Dutil 1997). There is a high degree of seasonal variation in condition (Yaragina and Marshall 2000). The timing of research surveys may not always be optimal for establishing a representative annual condition measure. Traditionally, weight information has not been collected on the IBTS. Consequently, the data available to describe interannual variation in condition is very limited. Figure 2.7 shows a reconstruction of relative condition factor (Fulton's $\mathrm{K}=$ weight/(length) ${ }^{3}$ ) for North Sea cod ages 1-6. The time period 1996-2001 suggests that there is rapid fluctuation that is not synchronous across age-classes. For cod ages 4 and 6 there is some suggestion of condition values being lower in 2001 compared to values for 1979 to 1981. Due to the short, incomplete nature of the time series it is not possible to determine whether there is a correlation with recruitment trends. However, it should be noted that other gadoid stocks show positive associations between condition and recruitment (Blanchard 2000, Marshall et al. 1999).

### 2.3.1.3 Female-only biomass

As a result of dimorphism in growth, mortality and behaviour, sex ratios shift towards a higher proportion of females with increasing size. The loss of large/old spawners affects the number of mature females disproportionately. This could result in losses of reproductive potential that are disproportionate to the loss of spawner biomass. For example, more of the variation in recruitment of Baltic cod can be explained using female-only biomass compared to the biomass of males and females combined. Female-only biomass for North Sea cod was calculated as the product of numbers at age, sex ratio-at-age, female maturity-at-age and weight-at-age. Numbers-at-age of cod were available from the North Sea Demersal Working Group report (ICES 2002a). Female maturity-at-age, sex ratio, and weight-at-age were obtained from the English groundfish surveys carried out in the North Sea during the third quarter. Only ten surveys had sufficiently resolved data. Therefore, estimates of female-only biomass were confined to the years shown in Table 2.2.

Table 2.2. Results of model fitting for North Sea cod.

| Year | Female-only SB <br> $($ metric t $)$ | VPA SB | Recruitment at <br> age 1 | Ricker fitted <br> recruitment | Residuals |
| :--- | :---: | ---: | :---: | :---: | :---: |
| 1979 | 56962 | 164268 | 899522 | 363008 | 536514 |
| 1980 | 87145 | 181869 | 314766 | 380027 | -65261 |
| 1981 | 101074 | 195732 | 618498 | 391356 | 227142 |
| 1989 | 47887 | 90611 | 133940 | 253088 | -119148 |
| 1990 | 25889 | 78040 | 168570 | 226865 | -58295 |
| 1996 | 27339 | 76358 | 421717 | 223165 | 198552 |
| 1997 | 31428 | 80187 | 69536 | 231519 | -161983 |
| 1998 | 11500 | 71553 | 139369 | 212343 | -72974 |
| 1999 | 27497 | 61467 | 215023 | 188356 | 26667 |

Within this time period there was a weak but insignificant ( $r^{2}=0.28, p=0.14, n=9$ ) linear relationship between female-only SB and recruitment (Figure 2.8b); however, the overall pattern of variation was similar to that observed in the conventional stock-recruitment relationship (Figure 2.8a). It should be noted that the maturity assessment in the third quarter underestimates the SB of the younger first-time spawners. There was no significant $\left(r^{2}=0.09, p=0.14, n\right.$ $=9$ ) relationship between female-only SB and the residuals from the S-R relationship.

### 2.3.2 Incorporating this information into management

The North Sea cod stock was initially considered data-poor in terms of analysis of reproductive potential. However, the analyses undertaken during this meeting of SGPRISM revealed that there were data resources which could be utilized (Section 2.3). The perception that there is no data is not justified! However, existing data often are discontinuous in time or subject to sampling biases. More definitive databases are required to describe the relationship between SB and reproductive potential and between reproductive potential and recruitment conclusively.

There are several considerations when reconstructing historical trends in reproductive potential or collecting new data:

- When reconstructing historical trends in reproductive potential for data-poor stocks, missing data are often extrapolated from another stock in spite of evidence of large inter-stock differences in the fecundity/length relationship (Oosthuizen and Daan 1974). Until more is known about the stock-specific nature of reproductive parameters and the degree of spatial and temporal variation in. sex ratio, maturity, fecundity, and condition, it is inappropriate to extrapolate this information between stocks.
- It is essential to collect time series of reproductive data if reproductive potential is to be successfully incorporated into stock assessment and projections. Emphasis must be given to the quality of surveys and appropriate survey design. Surveys must take place at times relevant to the reproductive process being considered, e.g., maturity should be assessed approaching the spawning period and condition prior to vitellogenesis. Surveys should also account for biases caused by spatial patterns, fish behaviour and sampling strategies such as stratified length sampling.
- In the case of North Sea cod data obtained from market sampling has limited value because gutted fish are landed (e.g., no condition or fecundity information). Consequently, data obtained from surveys are critical to characterizing the dynamics of growth and reproduction (see also Section 4.4.3).
- Maturity data should be collected and reported by sex and be of suitable resolution for the correct interpretation of diverse spawning strategies.
- In the absence of data, simulation approaches may be advantageous to examine the effect of variation of weights and maturities on medium-term stock projections.
- New data on fecundity and egg quality of North Sea cod may become available in the near future (Witthames et al. WD2).


## Data-rich stocks

$\underset{\text { constant } \mathrm{m}, \mathrm{w}}{\mathrm{SB}} \Rightarrow \underset{\text { variable } \mathrm{m}, \mathrm{w}}{\mathrm{SB}} \underset{$|  gender-specific  $\mathrm{m}, \mathrm{w}$ |
| :---: |$}{\text { Female-only } \mathrm{SB}} \Rightarrow \underset{\text { fecundity }}{\mathrm{TEP}} \Rightarrow \underset{\text { egg quality }}{\text { TVEP }} \Rightarrow \underset{\text { structure of spawning stock }}{\text { Spatio-temporal EP }}$ vs. Rec

quantitative advance if a qualitative advance, by definition $\longrightarrow \underset{\left.\begin{array}{c}\text { recruitment variation is }\end{array}\right]}{\begin{array}{l}\text { higher proportion of }\end{array}}$

## Data-moderate or data-poor stocks

Proxies for reproductive potential<br>Recruitment<br>e.g., age diversity indices, condition indices<br>vs.

Residuals from $\mathrm{S} / \mathrm{R}$ relationship
quantitative advance if a
higher proportion of
recruitment variation is explained

Figure 2.1. Schematic illustrating the differences between qualitative and quantitative advances in quantifying reproductive potential. For data-rich stocks qualitative advances are achieved by replacing spawner biomass (SB) to total egg production (TEP), total viable egg production (TVEP), and spatio-temporally disaggregated egg production (EP) through the addition of more highly resolved information on maturity ( m ), weight ( w ), sex ratio, fecundity ( f , egg quality, and spatio-temporal distribution of spawners. Quantitative advances result if a higher proportion of recruitment variability is explained by the new index. Data-moderate or datepoor stocks are limited to comparing proxies of reproductive potential either to recruitment or to residuals from the stock-recruitment relationship.


Figure 2.2. The stock/recruit relationship for Northeast Arctic cod using different indices for reproductive potential. a) spawner biomass calculated with constant values of proportion mature and weight at age for pre-survey time period (ICES 2001d); b) spawner biomass calculated with year-specific values of proportion mature and weight at obtained from Russian and Norwegian historical sources (ICES 2001b); and c) total egg production. The recruitment index used is the abundance at age 3. Observations are labelled by year. The loess curve (degree=1, span=1.5) is shown for each.


Figure 2.3. Time series of spawner biomass (solid line; from ICES 2001b) and total egg production (dashed line).


Figure 2.4. Length maturity ogives (estimated from logit transformations) of North Sea cod for males (solid symbols and females (open symbols), according to data by Holt (1893) (squares) Graham (1924) (diamonds) and Oosthuizen and Daan (1974) (circles). $\mathrm{L}_{50}$ values indicated by the date.


Figure 2.5. Interannual variation of age diversity of cod in the North Sea based on the Shannon diversity index. Dotted lines denote long-term means.


Figure 2.6. Recruitment in North Sea cod (age 1). A) against age diversity H (time invariant proportion mature); B) against SB in tonnes (1963-2000); C) against age diversity H (with variable proportion mature, open diamonds and fixed proportion mature, closed squares) (1980-1995, Cook et al. 1999, ICES 2002a); and D) against SB in tonnes (1980-1995). Linear relationships shown where applicable.


Figure 2.7. The time series of relative condition factor at ages 1 to 6 for the North Sea cod (combined sexes) calculated from the English groundfish survey data sets. The values are scaled in order to separate the time series so that the year effects can be distinguished.

A)

B)

Figure 2.8. A) The relationship between SB of North Sea cod and recruitment at age 1. Labels denote year; and B) the relationship between female-only SB and recruitment at age 1. Labels denote year.

### 3.1 Introduction

This section is intended to provide a brief summary of process studies in selected areas, touching on the potential for using the results from these studies in management advice. This is followed by a discussion focussing on general aspects of the use of process studies in this context.

### 3.2 Area case studies

The areas covered here represent the interests of members of this meeting of SGPRISM. They are also notable for the differing extents to which the areas have been covered by process studies, ranging from highly detailed and systematic coverage (the Baltic) to much sparser coverage (the North Sea).

### 3.2.1 Baltic Sea

The processes affecting recruitment to stocks in the Baltic Sea have been the subject of extensive studies under the EU CORE and STORE programmes. A general introduction to these programmes is given by MacKenzie and Köster (2001), and the following summary of progress in the STORE project, with particular reference to cod recruitment, is adapted from Schnack and Köster (2001).

The objectives of the STORE project are the following:
a) Determine stock-recruitment relationships for Baltic cod and sprat in relation to key environmental factors influencing the production of viable spawn and the survival of early life history stages.
b) Improve short-term predictions of stock development by integrating recruitment estimates based on the present status of the stock and its biotic and abiotic environment.
c) Develop predictive recruitment models for medium- to long-term forecasts of stock development under different environmental and fishery scenarios.
d) Estimate biological management reference points, critical stock limits and target spawning stock sizes based on stock-recruitment relationships and stock development simulation models, and considering the precautionary approach for fisheries management.

On the basis of extensive retrospective data compilation and verification, simple correlation analysis have first been made between SSB and stage I egg production derived from ichthyoplankton surveys and between 4 successive developmental stages of cod and sprat (Table 3.2.1). When concentrating on those sub-areas which were best sampled for the respective species, for cod there is a significant correlation between the abundance of larvae and 0 -group stage but not between any of the earlier stages. This indicates that variations in spawning stock structure and in environmental conditions during egg to larval development are critical for recruitment success of Baltic cod, whereas the larval to juvenile development appears to be more stable. In sprat the main regulation seems to take place in the larval to juvenile phase. In the following only cod is further considered with a focus on the critical early stages.

Table 3.2.1. Correlation coefficients for successive early life stages of cod in Sub-division 25 and sprat in Sub-division 26

| Variable 1 | Variable 2 | Cod <br> SD 25 | Sprat <br> SD 26 |
| :--- | :--- | :--- | :---: |
| SSB $^{\text {II }}$ | Egg production stage 1 | 0,19 | $\mathbf{0 , 6 6 *}^{*}$ |
| Egg production stage 1 | Egg production stage 3 | 0,51 | $\mathbf{0 , 8 2 *}$ |
| Egg production stage 3 | Laval abundance | 0,36 | $\mathbf{0 , 8 1 *}$ |
| Laval abundance | 0-group abundance | $\mathbf{0 , 8 0}{ }^{*}$ | 0,32 |

* correlation coefficients significant at $\mathrm{p}<0.01$
${ }^{1)}$ based on maturity ogives invariate over time

In Table 3.2.1, SSB is based on constant maturity ogives as used previously in standard stock assessment. When including the actual fluctuation in maturity ogives and using only the mature female stock component, this is already significantly correlated to stage I egg production. Moreover, relative fecundity changes over the years, with an obvious general trend to higher values in more recent years (Kraus et al. 2000). On population level, this fecundity appears to be clearly related to the food supply (Kraus et al. 2001), expressed as suitable clupeid (sprat age-groups 1+ and herring 12) biomass per predator weight. Temperature had an additional but very minor effect. SSB, thus, appears to be too
coarse a measure of the potential egg production; information on maturity and fecundity has to be considered in addition.

When calculating the potential seasonal egg production, based on area specific mature female stock components and relative fecundity values, predicted from food availability, this leads to a significant relationship with the observed daily egg production as obtained from egg surveys during peak spawning time. A corresponding relationship is obtained also with the observed total seasonal egg production, but can be based on a rather limited data set only.

Turning now to the second phase in the correlation scheme, which addresses the variability in egg mortality, for cod this phase is assumed to be largely depending on the oxygen conditions in the deep-water layers. To consider this factor the Recruitment Volume (RV) has been defined by Plikshs et al. (1993), which identifies the pelagic habitat volume in the depth range with sufficient salinity for fertilization success and floating of cod eggs, and with an oxygen level above the minimum required for egg development. This RV has frequently been utilized in stock-recruitment considerations, but unfortunately it explains only a limited part of the variation in egg mortality (Köster et al. 2001a). This can be related to the following problems: When using the total RV in the Baltic, combined for the different spawning areas, as a measure of oxygen related survival, a large and varying amount of the total egg production is placed in areas with no RV present, and also large and varying portions of the total RV may be available in areas with no egg production. Thus, individual spawning areas have to be considered separately. In addition, the proportion of the total egg production placed within the vertical range of the RV varies according to the varying salinity profile, and the survival of these well placed eggs is largely depending on the widely varying oxygen profile within the limits of the RV. So this vertical information has to be utilized as well.

From experimental work, the relative viable hatch has been related to oxygen and this relationship been utilized to define an improved oxygen related survival index (OES). This index is based on the observed oxygen profile and the predicted vertical egg distribution, derived from the salinity profile (Köster et al. 2001b). This new survival index follows the variation in RV only in the very general trend; for specific years or periods there are substantial differences. Predicted egg survival, based on the OES, is significantly related to observed daily egg survival as derived from successive egg surveys during peak spawning; and also when applying OES on the estimated potential egg production, then the predicted surviving seasonal egg production explains a significant part of the variation in the observed seasonal production of stage III eggs. This could not be achieved with the RV. The remaining variance is still high though, but this is not surprising as the variance between potential and realized egg production is still included and other factors, especially predation, are effective in addition to oxygen.

Quantitative estimates of predation on early life stages are difficult to obtain (Bailey and Houde, 1989). The predation by sprat is rather high in the early years of this time series but is much reduced in the more recent years, whereas predation by herring has increased, and in total the relative importance of predation has decreased. This shift in the importance of the two clupeid species can be explained by the shift in the peak spawning time of Baltic cod from spring into summer period. Sprat leave the central parts of the basins after spawning and feed in more shallow waters, and herring having spawned in the coastal areas return back into the central parts.

In addition to these changes in the horizontal overlap of predator and prey, variation is also observed in the vertical overlap, depending on the salinity and oxygen profiles. When combining predator abundance and vertical overlap of predator and prey, the results indicate an especially high predation on cod eggs in the second half of the 1980s and much lower levels in the years before and in recent years.

In the early larval phase, predation appears to be much less important (Köster and Möllmann 1997), but during this stage the food supply may be more relevant, especially when considering changes in the zooplankton composition, described by Möllmann et al. (2000, 2001). According to light conditions, visual performance and vertical distribution of the larvae, the food supply in the depth range of $25-50 \mathrm{~m}$ should be most critical, as described by Grønkjær and Wieland (1997). The seasonal distribution of abundance of nauplii has been modelled for the Bornholm basin (Hinrichsen et al. 2001a) and presented as average concentration in dependence of the total water depth. When all species are considered, then the concentration of nauplii is not substantially different between deep and shallow areas and a marked seasonal maximum is showing up in May-June. Excluding Pseudocalanus elongatus, a species which has reduced in abundance in more recent years (e.g., Mölmann et al. 2000), then the seasonal maximum of nauplii is shifted to the autumn and concentrations are much lower in the deeper parts of the basin, where the early larvae are mainly placed, as compared to the more shallow areas. This result shows the potential relevance of the observed changes in species composition of zooplankton in recent year for cod larvae survival.

For the larval to juvenile phase, mortality was indicated to be not as variable as for eggs (Table 3.2.1). Nevertheless, two factors appear to be of importance: the drift of the pelagic stages and the mortality due to cannibalism.

In the drift pattern of larvae, yearly differences may be quite striking as derived from hydrodynamic modelling (Hinrichsen et al. 2001b). The drift into shallow areas appears to be essential for the settlement of the early juvenile stages and successful further development. Accordingly a larval transport index has been developed, based on the wind conditions, to be included into the cod recruitment model.

The second factor, cannibalism, is indicated by MSVPA to be a major potential source of pre-recruit mortality (e.g., Neuenfeldt and Köster 2000). However, large differences were obtained in the suitability coefficients of 0 -group cod as prey obtained from MSVPA runs based on sub-sets of stomach content data covering different time periods. These differences were obvious only for predators above age 4, related obviously to a decline in spatial overlap with the juvenile stages at decreasing adult stock size (Uzars and Plikshs 2000). Thus, a substantial variability in pre-recruit mortality due to cannibalism is to be expected only if substantial amounts of older age classes are present in the stock.

Those factors identified as significant in the different spawning areas have been included into statistical stockrecruitment models, built up for each of the three major spawning areas separately. The effects were then combined for the Central Baltic in total. Factors that have been considered are: potential egg production, clupeid egg predation, oxygen content in the reproductive volume and larval transport index (the latter in Sub-division 25 only).

The combined model describes fairly well at least the general trends in recruitment variability as obtained from MSVPA. However, differences may still be substantial for individual years, e.g., 1979. In order to test its predictive power, the model was refitted again to the time series of data, excluding alternatively the first or the last 4 years. The models derived this way could reasonably well predict the recruitment success in these excluded years, again with exception of year 1979. Also for the most recent years after 1995, not included in any of the fitting procedures, the predictions compare well with the VPA estimates. This, however, is a less rigid test, as only minor changes occurred in stock size and recruitment during this last period.

In conclusion, the basis for prediction of Baltic cod recruitment has obviously been improved, but the model so far utilized is still a conventional statistical approach. The improvements have been achieved by using more of the available information and including newly provided results, which allow for instance to estimate the potential egg production more reliably as compared to using SSB as a too coarse proxy only. To a large extend improvements are related to a better spatial resolution of information: i.e., individual spawning areas have been handled separately, predator-prey overlap, larval transport and the distribution of eggs within the oxygen profile including consequences for survival have been considered. The remaining variance in the recruitment model may be due especially to parental effects on early life stages (Vallin and Nissling 2000) and to the prey availability for larvae (Hinrichsen et al. 2001a). Both factors have not yet been considered in the model but are a focus of study in the STORE project.

In addition to the spatial aspects, further improvement of resolution in time will be essential. So far, average environmental conditions in the main spawning times and areas have been considered. However, these may not always be the most relevant ones. Backtracking of actual survivors to their birth dates and places, based on age analysis and drift modelling, shows that they can be originated to a large and varying extend from outside the spatial and temporal centers of egg production (St. John et al. 2000).

Thus, describing characteristics of survivors and identifying those environmental conditions on adequate scales, relevant for growth and survival, is a main focus of the project. On the modelling side, purely statistical models, assuming linear though multifactorial relationships, do not seem to be satisfactory for predicting purpose. Thus, the construction of Individual Based Models (IBM) is attempted in order to follow the developmental success on the basis of a number of functional relationships and non-linear responses to environmental conditions. This is done in combination with hydrodynamic and also some food web modelling to provide the relevant environmental input data to the IBM. This approach introduces a better resolution in time and allows to study the sequential effects for changing conditions, which may be different from effects of constant average conditions. Also, it allows to include the variance in the performance of individuals within a cohort and its relation to characteristics of the parental stock.

## Potential use of process information in assessments for Baltic cod

As the above summary notes, the STORE project includes the intention to incorporate information obtained on recruitment processes into medium-term projections and reference points for cod and sprat. WD12 summarised some preliminary investigations into this for Baltic cod. This noted that most of the existing studies on stock and recruitment used multiple regression models, which were not suitable for direct use in stock projections. This was partly because regression models are not appropriate for use with values outside the range of observations, and partly because the studies summarised above have used multi-species assessments which differ in age range and area coverage to the stock units used in the routine assessments. However, these studies have highlighted that potential egg production and
reproductive volume are important determinants of recruitment, and future work will focus on the incorporation of these effects into stock projections.

### 3.2.2 Barents Sea

## General

It is believed that the state of the Barents Sea ecosystem to a large extent will be revealed through the state of the stocks of Northeast Arctic cod, Barents Sea capelin and Norwegian Spring-Spawning herring (Hamre 1994). Both total fish production in the Norwegian-Barents Sea area (including Norwegian coastal waters), and also other aspects of the total ecosystem, are believed to be closely linked to the development of these stocks. Cod prey on capelin, herring and young cod (Bogstad and Mehl 1997), while herring is an important predator on capelin larvae (Huse and Toresen 2000). Cod growth is also affected by prey (especially capelin) abundance (Mehl and Sunnanå 1991). All three stocks show large variations in recruitment.

In the present case study on including process information in stock-recruitment models for fish species in the Barents Sea, we will thus only consider these three species. Appropriate modelling of other processes (growth, maturation, predation/species interactions) are also important both for short-term and medium/long-term stock prognosis for these three stocks. However, in this context we will focus on how to incorporate process information in stock-recruitment models for these three species.

## Cod

The recruitment of Northeast Arctic cod is quite variable, and several processes and variables have been suggested as explanation: Spawning stock biomass, egg production, temperature in the spawning/larvae period, NAO, cod cannibalism, prey abundance, wind and currents etc. (see e.g., Marshall et al. (2001,WD 9), Ottersen and Sundby, 1995; Ottersen and Stenseth, 2001; Bogstad et al., 1994). The issue of replacing spawning stock biomass with other measures (e.g., total egg production) of reproductive potential has been elaborated on in Section 2.

For this stock, there is a need for predicting year-classes, which have been spawned but have not yet entered the fishery, in order to improve short-term predictions. Also, there is a need for incorporating process information to improve medium- and long-term predictions. These two issues will be addressed separately below.

Northeast arctic cod enters the fishery at age 3. The ICES Arctic Fisheries WG has for many years (since 1987) used the ICES computer program RCT3 (formerly RCRTINX2) (Shepherd, 1997) to predict the strength of recruiting year classes at age 3. This is a simple calibration and regression method, which combines multiple time series of recruitment indices and VPA estimates of year-class strength at the age of recruitment to the fishery. It does not take into account at which life stage an index is measured. Also, RCT3 does not allow for including information on environmental conditions, cod cannibalism and other processes, which are important in determining recruitment of this cod stock. A model that predicts the abundance of NEA cod of a cohort of age 3 based on available information about the abundance at earlier life stages, as well as process information, should be constructed. Tretyak (1999) constructed a model of the mortality from ages $0-3$ which is of this kind, but there is a need for a model incorporating both Norwegian and Russian data sources and recent knowledge on process information (see WD 5 for a summary of some recent studies).

For medium- and long-term predictions, a population model for cod where the recruitment is modelled accounting for fecundity information and cannibalism could be a reasonable first step to aim for. Thereafter, the biological realism of the predictions could probably easier be improved by using a cod-capelin-herring model rather than by including more process information in the stock-recruitment model for cod. Cod cannibalism is inversely related to capelin abundance (ICES 2001b), which could be accounted for using a multispecies model.

For this stock, it is also important to model the length of the age 3 recruits, as this has been observed to vary from 28 to 41 cm , according to Norwegian survey data (ICES 2001b).

## Herring

Norwegian spring-spawning herring spawn along the Norwegian coast, but abundant year-classes are generally found in the Barents Sea approximately from age 6 months until age 3. Year-class strength seem to be determined in first half year of life; the larval survey in April is a good indicator of spawning stock size but useless for predicting recruitment, while the 0 -group survey in August-September is a reasonable predictor of recruitment. Temperature variations are also
an important factor affecting recruitment, but according to Toresen and Østvedt (2000) it was heavy fishing pressure and not environmental variability which was the main cause of the collapse of this stock in the late 1960s.

The recruitment model used by WGNPBW (ICES, 2001c) is a traditional Beverton-Holt recruitment model where the parameters are estimated on a log scale. However, the recruitment is highly dynamic with a few outstanding year classes. To better adapt the model to this stock $10 \%$ of the highest recruitments are excluded from the regression. When a draw from the recruitment model is made these year classes are selected with a $10 \%$ probability and a draw with equal probability is made. If the highest recruitments are not selected the recruitment is given as the exponentiation of the logarithm of the Beverton-Holt model with a random draw from the residuals added. This relationship does not contain any process information.

As for cod, there is a need for including process information in prediction of herring year-classes that already have been spawned (from age 0 to age 3 at which the fishery starts). However, much less survey information is available for age $0-3$ herring than what is the case for cod.

## Capelin

Predation by juvenile herring on capelin larvae seems to have a strong impact on the recruitment to the Barents Sea capelin stock. Hamre (1985) was the first author to suggest that the decline in the capelin stock could possibly be associated with the presence of young herring in the nursery area of the capelin larvae. In years with high abundance of juvenile herring in the Barents Sea, capelin recruitment is severely hampered (Gjøsæter and Bogstad 1998).

Capelin spawn approximately 1 April. In 'herring' years the abundance of capelin larvae, as measured during a larval survey in June, seems to be adequate for securing recruitment. However, in these years nearly all 0 -group capelin disappear between this June survey and the 0-group survey (Anon. 1997) in August-September (WD 1). Predation by herring on capelin larvae has also been observed (Huse and Toresen 2000). The effect of herring on capelin recruitment also seems to be linked to the overlap between juvenile herring and 0-group capelin (Gjøsæter and Bogstad 1998).

Gjøsæter and Bogstad (1998) fitted the following curve to the stock-recruitment data for capelin and abundance data for young herring:
$R=\frac{758 S S B}{S S B+74+2797 H_{1+}}$, where $\mathrm{H}_{1}+$ is the biomass (million tonnes) of one year and older herring in the Barents
Sea from the acoustic young herring surveys in the Barents Sea during May-June. This model explains $87 \%$ of the variation in recruitment at age 1 . The relationship should be revised based on new data on herring abundance and capelin spawning stock biomass, but is formulated in such a way that it can be utilised for short- and medium-term predictions of the capelin stock. The abundance of young (1+) herring in the Barents Sea can be predicted reasonably well one year ahead, based on survey data.

Barents Sea capelin has almost total spawning mortality, and is managed by a target escapement strategy (Gjøsæter et al., 2002). The management is based on an acoustic survey in September, which is considered to give an absolute abundance estimate of the abundance of $1+$ capelin. A length-dependent maturation model is then used to determine the maturing stock, which is predicted half a year ahead based on a multispecies model where predation by cod is taken into account. The fishery is conducted in January-April, mainly on the spawning stock. It has been discussed in the Northern Pelagic and Blue Whiting Fisheries Working Group (ICES, 2001c; Anon., 2001a) how the target escapement strategy could be made dependent on the expected abundance of young herring in the Barents Sea in the year of spawning, but no method to do this has yet been found.

## Prediction models which could be used

The multispecies model Bifrost (Anon., 2001a) could be used for making predictions for the cod-capelin-herring system in the Barents Sea. This model is an age-structured multi-species model where predation, growth, recruitment and fishery are modelled. It is presently used for investigating harvest control rules for Barents Sea capelin.

One way of including the process information on reproductive potential (Section 2) is to use a population model with an age and length structure. For Northeast Arctic cod, such a tool is available - Fleksibest (Frøysa et al., 2002). Fleksibest is a single-species application to Northeast Arctic cod of the multi-area, multi-fleet, multi-species age-length structured model framework Gadget (formerly BORMICON) (Anon., 2001b). Here, the population dynamics processes growth, maturation, predation and fishing are modelled as length-dependent processes. A stock can be divided into a mature and
an immature part, as is done in Fleksibest. It is easy to include also information on sex ratio and on fecundity, as well as letting the condition factor influence the processes in addition to the length. Gadget currently lacks a stock-recruitment module (recruitment is estimated or read from file), but there are plans to include such a module. Using such a flexible framework allows for studies of the effect of changes in growth, exploitation rate, and selection pattern on stock development.

### 3.2.3 North Sea

The only species to be examined for the North Sea by the Study Group was herring (Clupea harengus) (WD6). The basis of the work presented was an interest in exploring life history trajectories in an attempt to understand the factors that determine year class strength in fish populations. In general, there are very few species where there is sufficient information to undertake relatively detailed analyses of the different life history stages.

One exception is the long series of survey indices for North Sea herring. Fishery independent data are available on the age, length, mean weight, fecundity and abundance of the adults, larvae, O-ringer pelagic stage, 1-ringers and 2-ringers. Thus providing information on estimated number of eggs and abundance at spawning, 0-3 months, 3-6 months, 1 year and 2 years within the life of a cohort. The data used in this study are generated from survey data and are assumed to be linearly related to the abundance of the life stage measured. Since there are no estimates of absolute abundance the relationships between life stages can only be considered as relative.

There is an effect of age structure of the adult population on the numbers of eggs produced. There are relationships between egg production and the numbers of larvae but there is a suggestion that the relationship has shifted over time (a regime shift in the North Sea during the late 1980s (Reid et al. 2001)). In general the numbers of early juveniles are positively correlated with the numbers of larvae, however, there are some notable exceptions. There are a series of three years (1988-1990) plus 1997 when there was poor survival from larvae to early juveniles and one year (1986) when survival appeared particularly good. In the next transition from early juveniles to 1-ringers there was a good correlation between years with one exception (1995). This was a particularly strong year class. In the transition from 1 to 2 -ringers there was again correlations between successive years, however, there appeared to be a shift in survivorship for the periods pre and post 1990. These data are visualised in the form of Paulik diagrams (Paulik 1973), four panel graphs (Figure 3.2.3.1, note that neither the estimated egg production nor 2-ring abundance are shown in this figure) that highlight where in the life history changes in correlations between years in each year-class occur.

It is apparent that good or bad year classes are formulated at different time periods within life history. Variation in abundance of potential predators and/or competitors were examined in conjunction with the abundance of herring at the different life history stages but these did not shed any light on why there are apparent shifts in survivorship. Without detailed studies on the predator/prey relationships for herring it is not possible to try and determine the processes, which are determining year class strength.

In reviewing stocks within the North Sea it became very apparent that there are no other stocks with sufficient information through the life history to undertake studies such as these. In the same vein there was not sufficient information on North Sea cod to undertake similar studies to those undertaken on Baltic cod and sprat or Barents Sea cod and capelin. Part of this may stem from the more complicated species interactions seen in the North Sea compared with either the Baltic or Barents Seas.


Figure 3.2.3.1. Paulik diagram for North Sea herring using survey data. Axes show the units for each index and are not absolute values. SSB estimated from acoustic surveys, Mlai (Multiplicative Larvae abundance Index), MIK 0 -ring (Methot-Isaacs Kidd 0 -ring index) and 1-ring IBTS (1-ring index from the International Bottom Trawl Surveys). The SSB/Mlai panel has data divided into pre and post North Sea 'regime shift' (see Reid et al. 2001). In the Mlai/MIK panel the years shown as triangles have been removed from the regression. In the MIK/1-ring panel the 1995 data point (open diamond) has been removed from the regression.

### 3.3 The management context for process information

Management advice for an ICES stock will typically be based on an age-based assessment which estimates the current state of the stock, which is then compared with pre-determined reference points. A short-term catch forecast will be used to present catch options for the forthcoming year for a range of possible levels of fishing mortality, and mediumterm projections will be used to make a statement about the probability of a given management objective being achieved as a result of the various management options which are presented. In the absence of explicit management objectives for most stocks, these statements are usually framed around the probability of returning or maintaining the stock above specified biomass reference points.

This system presents two areas where process information might be incorporated: in the determination of reference points, and in the projections used to estimate the probability that the management targets are achieved in the mediumterm.

### 3.3.1 Precautionary Approach reference points

The precautionary approach to fishery management provides the framework for the fishery management advice provided by the ICES Advisory Committee on Fishery Management [ACFM] (ICES, 2000b). This states that reference points will be stated in terms of biomass and fishing mortality rate. These two indicators can be divided into one describing an attribute of the resource (biomass) and one describing an attribute of the fishing fleet (fishing mortality). While the two indicators are related, it is clear that process information is most relevant to the resource indicator. The use of the two indicator scales is summarised in the following extract from the ACFM advice:

In order for stocks and fisheries exploiting them to be within safe biological limits, there should be a high probability that 1) the spawning stock biomass is above the threshold where recruitment is impaired, and 2) the fishing mortality is below that which will drive the spawning stock to the biomass threshold, which must be avoided. The biomass threshold is defined as $B_{\text {lim }}$ (lim stands for limit) and the fishing mortality threshold as $F_{\text {lim }}$.

From this extract, it can be inferred that the ICES implementation of the precautionary approach is framed around a rather simple model of stock dynamics, i.e., that there is a specific value of spawning biomass below which recruitment is impaired, and that fishing mortality is the only external factor which influences the size of the spawning stock. Such a simple approach is open to criticism, particularly in view of the information becoming available from process studies. Before elaborating on these criticisms however, it should be noted that the simplicity of this conceptual model has a number of advantages from a management perspective. Firstly, the two indicators, SSB and mean fishing mortality, are readily available from age-based stock assessments, so their estimation does not require data or modelling beyond what is already routine. Secondly, the simplicity also implies a correspondingly simple control rule for management action, e.g., 'If $\mathrm{B}<\mathrm{B}_{\text {lim }}$, then Take Action'. Thirdly, the implication that SSB is influenced only by fishing mortality is often not far from the truth for heavily exploited stocks, with the added justification that fishing mortality is usually the only factor influencing SSB which fishery managers can seek to manage.

As noted above, the results from various process studies imply a number of potential criticisms of the conceptual model underlying the ICES implementation of the precautionary approach. Firstly there is the implicit assumption that SSB is an adequate measure of the stock's reproductive potential, when there is an increasing amount of evidence to indicate that this is not the case - see Section 2. Secondly, there is the assumption that there is a fixed level of SSB (or reproductive potential) below which recruitment is impaired. This in turn implies that recruitment is determined only by reproductive potential; another point which is not consistent with the results of process studies. These points also have implications for the effectiveness of fleet reference points as in cases where stock size is not the only important influence on recruitment, a greater reduction in fishing mortality may be required to ensure a given probability of stock recovery. This in turn implies that process studies have implications for both resource (biomass) reference points, and fleet (fishing mortality) reference points.

Within the context of the current ICES implementation of the precautionary approach, a number of approaches based on use of process information may be possible, for example:

- Use resource reference points which incorporate more biological complexity than SSB. At its simplest this could be no more than specifying a higher minimum SSB in cases where a high proportion of the spawning stock consisted of first time spawners. More complex measures, perhaps even incorporating environmental information might also be possible.
- Set reference points more conservatively in cases where effects other than fishing have a strong influence on the stock, in order that the additional uncertainty due to these other components is accounted for

Although they do not incorporate process information, it is useful to note that the segmented-regression models used by O'Brien and Maxwell (WD3 \& WD4) are an objective means of fitting a model which corresponds to the conceptual model behind the ICES implementation of the precautionary approach

### 3.3.2 Medium-term projections

A key feature of current medium-term projections is the use of a stock-recruitment model to estimate future recruitment. Fitting a stock-recruitment model in this way treats the other factors contributing to recruitment variation (apart from SSB ) as residual variation, which is assumed to be random, and using such a model in a projection implicitly assumes that this variation will continue to be purely random. A process study will, by nature, seek to identify other components of this variation. Hence the contribution that a process study may make to medium-term projections will be to increase the proportion of variance allocated to modelled effects and hence to reduce the residual variance. This will not reduce the variance around the projected recruitment estimates unless the additional factors are readily predictable. This applies whether the additional factors are biological, (e.g., fecundity), or environmental, (e.g., temperature). The uncertainty around projections may actually be increased by explicitly accounting for the variation of these additional parameters but this may actually be more realistic as current medium-term projection methods have been shown to under-estimate the uncertainty around stock projections (Patterson et al., 2000).

The reallocation of variance through the incorporation of process information represents a potential improvement to the procedure as it makes the assumptions about future variation of the additional factors explicit rather than implicit. This would allow presentation of medium-term projections for a number of scenarios; for example environmental change, as well as improvement of estimation of uncertainty, and thus also in the quality of probabilistic statements which are based on the medium-term projections. This point was addressed in the Report of the first meeting of this Study Group (ICES, 2000a), and some progress in these areas is described in Sections 5.1 and 6.1.

### 3.3.3 Practical implications of the incorporation of process information in assessment and advice

Process studies can give us the ability to understand why a stock produces a certain level of recruitment for a given number of individuals in that stock and how the stock will behave over space and time. The implication is that more realistic population models and stock projection models can be formulated.

In the context of stock and recruitment the process studies provide information and relationships, often relatively complex, on the factors that influence production and survival. In essence there is a possibility of providing new relationships between stock and recruitment that have a basis in biological processes.

The general problem is that at present very few process studies have been directly undertaken with the intention of input to the assessment and management process. By nature process studies are aimed at understanding what has happened in the past, and hence the models they use are detailed and descriptive, and are often based on multiple-regression-type approaches. These models are not usually suitable for use in projections, partly because of the need to forecast a large number of variables, and partly because of undesirable properties such as the possibility of generating negative recruitment or of generating a finite recruitment at zero stock size. Other drawbacks of these approaches are discussed in ICES (2000a).

The fact that process studies tend not to produce recruitment models which can be used directly within projections means that additional work is required to develop recruitment models for projections using the results from process studies. How this is done depends on the nature and results of the process study, but in general such work might involve identifying a small number of major factors, and considering how these might be incorporated in a recruitment model of an appropriate functional form, and how they might be projected forward. Similar considerations also apply to how external factors might influence such processes as growth and fecundity.

### 3.3.4 Consequences of incorporation of process information in management advice

As noted above (Section 3.3.1), the current ICES implementation of the precautionary approach assumes that SSB is an adequate measure of the reproductive potential of a stock, and that fishing mortality is the major factor influencing SSB. Under this model, stock recovery is assured if managers can reduce fishing mortality and hence allow SSB to return to above the biomass reference point. The simplicity of this conceptual model has advantages as it is readily understood by other stakeholders in the fishery. Introducing additional biological realism to the management advice would remove this advantage but would hopefully improve the quality and credibility of the advice by more explicit acknowledgement of the influences of effects other than fishing.

## 4 EFFECTS OF ENVIRONMENTAL VARIABILITY ON FISH STOCKS

### 4.1 Introduction

It is now well accepted that changes in the environment do impact fish population dynamics. This is evidenced both by the spatial scales over which recruitment appears synchronised for several stocks (Myers et al., 1995; Myers et al., 1997) and by apparent correlations between recruitment and environmental variables (Ottersen and Sundby, 1995; Myers, 1998; Planque and Frédou, 1999; Fox et al., 2000). A significant problem with advancing this work is that variables such as temperature are correlated with many other environmental variables which also affect fish at the individual (and by inference) at the population level. For example, sea surface temperatures in NE Atlantic during the first six months of the year are correlated with atmospheric conditions (NAO) and thus with wind stress and average direction. Wind stress in turn can affect primary productivity through water column processes whilst varying levels of turbulence have been linked with changes in feeding success of fish larvae (Alcaraz, 1997; Dower et al., 1997; Fiksen et al., 1998). Environment also affects the growth of adult fish both directly and through links with prey availability (Brown et al., 1989; Brander, 1995; Dutil et al., 1999; Ottersen and Loeng, 2000; Purchase and Brown, 2001). For mature fish, feeding conditions affect condition and the amount of energy which can be put into reproduction. Potentially, combined with whether the fish are first-time or repeat spawners, this could affect subsequent recruitment success (Marshall and Frank, 1999; Marshall et al., 1998). Changes in environmental variables such as temperature can also directly affect fish distribution and may lead to changes in spatial overlap with prey (Sparholt et al., 1991; deYoung and Rose, 1993; Heessen, 1993; Swain and Wade, 1993; Ottersen et al., 1998).

Models developed to explore such relationships vary from simple statistical approaches (examples above) through single species population dynamics models to complex, coupled hydrodynamic-biological models. To date the last category have tended to focus on the biology of early life history stages (Hermann et al., 1996; Leising and Franks, 1999). Their complexity and subsequent computer resource requirements currently prevent their application over
multiple years which is a requirement for extended time-series population dynamics modelling (Hermann et al., 2001). Similarly it can prove prohibitive to undertake multiple model runs for sensitivity testing although applications of engineering simulation theory can help in this respect (Megrey and Hinckley, 2001). Extended time-series population dynamics modelling has thus tended to use simpler, single species population dynamics models but representations of at least some biological processes can be included. The problem is usually knowing how much biological realism it is necessary to include to generate realistic results. The most useful aspect of such models is that they allow explorations of the sensitivity of the population dynamics to changes in rates and processes at different life history stages over several decades of simulated time (Quinlan and Crowder, 1999). Modern computers allow large numbers of model runs to be made producing the scope to explore the effects of stochastic variability. The challenge for the future appears to be in producing models capable of bridging these extremes.

In this section of the working group report we firstly consider some results from a simple, single species population dynamics model for cod. This model is used to explore how North Sea cod population dynamics might be affected by sea temperature. The model is used to assess the relative impacts of temperature on stock dynamics through affects on recruitment (early life history stages) and adult growth. The working question addressed is whether reduced recruitment linked with increased sea temperatures can be compensated for by increases in adult growth rates? One assumption of the model used is that condition (i.e., the relation between length and weight) is constant. In the second section we examine survey based and market sampling data available for North Sea cod and plaice for changes in adult condition factors. We then conduct a simple exploratory analysis to assess whether changes in such factors could be modelled using simple environmental indices as co-variates.

### 4.2 Modelling the effects of environmental change on North Sea cod

### 4.2.1 Introduction

Interannual changes in North Sea cod recruitment have been shown to be related to changes in sea surface temperature (SST) (Brander, 1996; Planque and Fox, 1998; Planque and Frédou, 1999; O'Brien et al., 2000). The relationship is negative with increased February to June SSTs resulting in reduced recruitment to the stock. Fluctuations in water temperature also have a secondary, continuous effect on cod individuals, through their influence of growth rates and hence weight at age. Cod stocks located in regions with low water temperatures (e.g., Labrador) generally have reduced rates of growth and reach a larger maximum size compared to fish stocks located in warmer waters (e.g., Celtic Sea) (Taylor, 1956; Brander, 1995)

Given this, the expected changes in climate and sea temperatures over future decades will ultimately influence the abundance and biomass of cod stocks. However, although increased temperatures may result in a reduced level of recruitment, this may be offset through higher rates of growth leading to higher Spawning Stock Biomass. We present the results of a study which investigates the combined effects of temperature on recruitment and weight at age, through the use of a self-generating model of North Sea cod.

### 4.2.2 Model functioning

## Overview

The model is a self-generating cod model (similar to WGMTERM (Reeves and Cook, 1994)), but with sea surface temperatures (SST) influencing the interannual changes in the numbers of recruits each January and bottom temperature influencing the changes in the mean length and weight of each age class (Figure 4.1). Abundances of new recruits (at age 1) are calculated using a combined Ricker-temperature stock-recruitment function which incorporates the previous Feb-May SST, and the previous years SSB. For each year, the abundance of age classes 2 to 15 are calculated by considering the levels of fishing and natural mortality (which for 1963 to 2000 were set at ICES 2001 WGNSSK values (ICES, 2002a)). Each month the average increments in length of each year-class are computed using a temperature dependent growth equation. These lengths are then converted into weight using a constant length: weight conversion (currently taken from FISHBASE), multiplied by maturity ogives (currently taken from ICES 2001 WGNSSK) and summated to give SSB of each age class are calculated. The main functional relationships in the model are as follows:

## Ricker-Temperature stock-recruitment function

This is an adaptation of the standard Ricker stock-recruitment function with temperature as an added parameter:
$R=S S B \cdot e^{(\mathrm{a}-\mathrm{b} . \mathrm{SSB})} e^{(c . S S \mathrm{~T})}$


Figure 4.1: Model interrelationships.
where $\mathrm{R}=$ the abundance of new recruits, $\mathrm{SSB}=$ Spawning Stock Biomass, $\mathrm{SST}=$ Sea Surface Temperature. Constants are statistically fitted parameters calculated using data from the $1963-2000$ period $(a=5.27542, b=0.00000355342, \mathrm{c}=-$ 0.560823 ).

## Negative Exponential Survival

This function calculates the number of survivors relative to the previous month given the natural and fishing mortality
$N_{t}=N_{t-1} e^{-(M+F) t}$
pressure upon the stock.
where $\mathrm{N}_{\mathrm{t}}=$ abundance of fish during that particular month, $\mathrm{N}_{\mathrm{t}-1}=$ abundance of fish during the previous month. $\mathrm{M}=$ instantaneous natural mortality rate. $F=$ instantaneous fishing mortality rate. $t=$ change in time in years (i.e., as monthly $\mathrm{t}=1 / 12$ ).

## Temperature influenced von Bertalanffy size at age function

The mean length of a particular age class may be calculated using the standard von Bertalanffy function given below:

$$
L_{a}=L_{\infty}\left(1-e^{-K(a-\text { tzero })}\right)
$$

where $\mathrm{L}_{\mathrm{a}}=$ length at age, $\mathrm{L}_{\infty}=$ asymptotic length (set at 119 cm ), $\mathrm{K}=$ growth parameter, $\mathrm{a}=$ age, tzero $=$ age where model registers that length is zero.

Work by Taylor (1956) examining differences in growth rates of cod stocks across the North Atlantic observed that temperature was related to both the rate of growth $(\mathrm{K})$ and the maximum length $\left(\mathrm{L}_{\infty}\right)$. Thus, K may be calculated as follows:
$\log _{10} K=1.22\left(\log _{10} T\right)-1.72$
where $\mathrm{K}=$ growth parameter, $\mathrm{T}=$ Bottom temperature .

Taylor (1956) obtained his parameters by relating the growth rates of many North Atlantic cod stocks with the mean water temperatures that these different stocks inhabit (similar models have been formulated by Brander (1995). However, here we are trying to relate relatively small changes in growth rates to similarly small changes in temperature over the 1963 to 2000 period. As such, the parameters in the above equation have been tuned ad hoc from those given in the Taylor paper, so that the model will produce weights with a similar mean to that observed over the 1963 to 2000 (see Figure 4.2.3.1.). Taylor (1956) also related differences in $\mathrm{L}_{\mathrm{inf}}$ to water temperature. However, it has been considered that this is not appropriate for this study when we are only dealing with changes in a single stock (hence $\mathrm{L}_{\text {inf }}$ is set at 119 cm (Gulland, 1977)).

For the calculation of the length of age 1 cod, the von Bertalanffy equation is used, with the mean annual bottom temperature (obtained from ICES) experienced during the first year of growth used to calculate K. For the length at ages $2+$ the differential form of the von Bertalanffy growth equation is used, which calculates the monthly change in length according to age \& K, and adds this to the previous length (Gamito, 1998):
$\frac{\partial l}{\partial t}=\frac{K}{12}\left(L_{\infty} .1-K . L_{t-1}\right)$
where $\mathrm{L}_{\mathrm{t}-1}$ is the length at the previous time step (in this case the previous month).

### 4.2.3 Evaluation of model performance and diagnostics

## Comparison of observed and modelled cod recruitment and SSB

The fit of the model was evaluated by producing output for the 1963 to 2000 period both incorporating and omitting the influence of temperature on the calculation of recruitment and SSB according to the following scenarios:

1) Ricker. Standard growth
2) Ricker + SST. Standard growth
3) Ricker. Temperature growth
4) Ricker + SST. Temperature growth

The output from each of the scenarios was then compared visually (Figures 4.2 to 4.5 ) and statistically (Table 4.1) against those values of recruitment and SSB for the 1963 to 2000 period given in the ICES WGNSSK 2001 report.

In order to enable the identification of how well the calculations of recruitment and growth perform separately within the model, two additional scenarios involved the replacement of modelled recruitment and growth with observed values taken from the ICES WGNSSK 2001 report as follows:
5) Observed recruitment. Standard growth
6) Observed recruitment. Temperature growth

Table 4.1. Sum of squares error $\left(\times 10^{-8}\right)$ between the observed and modelled recruitment and SSB.

| Model run | Recruitment | SSB |
| :--- | :---: | ---: |
| Observed recruits, Standard growth | 0.00 | 116.65 |
| Observed recruits, Temperature growth model | 0.00 | 185.07 |
| Ricker, observed weight at age | 22763.36 | 1736.60 |
| Ricker, Standard growth model | 20295.30 | 1211.87 |
| Ricker, Temperature growth model | 20193.57 | 799.69 |
| Ricker + SST, observed weight at age | 10384.32 | 790.86 |
| Ricker + SST, Standard growth model | 10916.53 | 1266.19 |
| Ricker + SST, Temperature growth model | 10309.60 | 415.64 |

The inclusion of SST in the calculation of the abundance of recruits notably improves the fit of the model to predict the abundance of recruits (Figures 4.2 to 4.5, Table 4.1). However, the additional inclusion of temperature information in the calculation of length (and thus weight) at age produces little additional improvement in the prediction of recruitment (c.f. Figure 4.2 with 4.4, and Figure 4.3 with 4.5), due to the negligible effect that changes in SSB (over the range of values considered here) are predicted to have on recruitment i.e., over this range of SSB the Ricker function is almost flat.

A greater improvement is observed when temperature information is incorporated into the prediction of SSB, as omitting the temperature information causes the model to overestimate the SSB. However, not too much credence can be placed on this, as the equation of Taylor (1956) has been tuned to mimic the observed weights at age. Unsurprisingly therefore, there is a reasonable similarity between the modelled and observed mean weights at age. Figure 4.6 shows the median modelled weight at age is too low for age 1 , slightly below the observed for ages 2, 5, 6, 7, 9 and similar for ages $3,4,8,10,11$. The variation in weight is reasonably captured by the model for ages 1 to 5 . However, older age classes have increasing inter-annual variation in weight which the model does not reproduce.

The greater variation in observed weight at age relative to modelled weight at age may be seen in Figure 4.6 (left hand plot). A comparison of the weight at age by year using correlation analysis (Table 4.3) shows evidence of similar trends in modelled and observed weight for age classes 1 and 2 . However, it is proposed that in the near future, the weight at age portion of the model will be more thoroughly reviewed and re-parameterised (see below).



Figure 4.2. Comparison of model output produced using the standard Ricker function and standard von Bertalanffy growth curve, with ICES values.


Figure 4.3. Comparison of model output produced using the temperature-Ricker function and standard von Bertalanffy growth curve, with ICES values.


Figure 4.4. Comparison of model output produced using the standard Ricker function and temperature influenced growth function, with ICES values


Figure 4.5. Comparison of model output produced using the temperature-Ricker function and temperature influenced growth function, with ICES values.

The distributions of observed and modelled weights at age were compared over the 1963 to 1997 period (Figure 4.6). A second comparison involved evaluating the observed changes in weight at age with time against the modelled weight at age with time (Figure 4.6).


Figure 4.6. Comparison of modelled (m) weights at age with observed values from working group report (w). Solid block part of bar represents $25^{\text {th }}$ to $75^{\text {th }}$ percentile, white horizontal bar the median.

The greater variation in observed weight at age relative to modelled weight at age may be seen in Figure 4.7 (upper plot). A comparison of the weight at age by year using correlation analysis (Table 4.2) shows evidence of similar trends in modelled and observed weight for age classes 1 and 2.



Figure 4.7. Comparison of observed weight-at-age 1963-1997 (upper plot), with modelled weight-at-age (lower plot) 1963-2000.

Table 4.2. Results of correlation analysis between observed and modelled interannual fluctuations in weight.

| Age class r | p |  |
| ---: | ---: | ---: |
| $\mathbf{1}$ | $\mathbf{0 . 5 3 6}$ | $\mathbf{0 . 0 0 1}$ |
| 2 | $\mathbf{0 . 3 8 4}$ | $\mathbf{0 . 0 2 3}$ |
| 3 | 0.209 | 0.228 |
| 4 | 0.326 | 0.056 |
| 5 | 0.307 | 0.073 |
| 6 | 0.122 | 0.484 |
| 7 | 0.086 | 0.625 |
| 8 | 0.153 | 0.38 |
| 9 | 0.203 | 0.241 |
| 10 | 0.037 | 0.833 |
| 11 | 0.016 | 0.927 |

## Examination of bias in modelled SSB

An examination of Figures 4.2 to 4.3 show that there is a bias in the modelled SSB, with the observed SSB being lower than the modelled SSB over the late 1960s/early 1970s. This is despite the model underestimating the weights at age (see above). In order to ensure that this bias was not caused by an inherent error in the calculation of recruits, residuals obtained from the modelled and observed recruits and SSB were plotted. In addition, a comparison of the stockrecruitment functions fitted to the data, with that used in the ICES WGNSSK report was made.


Figure 4.8. Residuals from modelled recruitment (left) and SSB (right) calculated using the temperature Ricker function.


Figure 4.9. Residuals from modelled recruitment (left) and SSB (right) calculated using the standard-Ricker function.

The residual plots (Figures 4.8 and 4.9), taken together with Figures 4.2 to 4.5 suggest that any bias in the SSB is simply due to the different abundances obtained in the calculation of the abundances of recruits. For example, the observed overestimation of recruitment in both the standard and temperature models from 1963 to 1970, and the
resulting increased cod abundances resulting from this for the next few years, is the cause of the later bias in SSB during the late 1960s/early 1970s.


Figure 4.10. Comparison of fitted stock-recruitment functions from the model (Ricker (no SST) and Ricker + SST with that in the ICES WGNSSK report (WG Shepherd). Note that the functions have not been bias corrected.

A final examination of the modelled stock-recruitment functions (Figure 4.10) suggests that there is little difference between the Ricker (no SST) and the WG Shepherd fits, except at the higher values of spawning stock biomass.

### 4.2.4 Conclusions

Models of this type can be used for examining the relative impacts of changes in environmental variables such as temperature on different life history stages of fish population dynamics. The model described above will be used in conjunction with sea temperature projections from coupled ocean-atmosphere circulation models to examine possible impacts of climate change on cod.

### 4.3 Modelling variation in North Sea cod condition factors

Weight and length at age data for individual fish are available from the English third quarter groundfish survey for the years 1979-1981, 1989, 1990, and 1996-2001. In addition, for each observation the position at which it was caught (latitude, longitude), the sex and age of the fish were available. Due to the scarcity of data for older ages and age 0 , only data for ages $1-5$ was used in the fitting of the linear models.

In order to examine the variation in Fulton's condition factor (weight/(length) ${ }^{3}$ ), taken to be a proxy for the quality of the spawning potential of the fish, a log weight - log length linear model was fitted to the data and the effects of explanatory variables on the variability about this relationship examined. The form of the fitted model was:

$$
\log (\text { weight })=\alpha+\beta \log (\text { length })+\delta \mathrm{A}+\varepsilon \mathrm{B}+\gamma \mathrm{C}
$$

where $\beta, \delta, \varepsilon, \gamma$ are the regression coefficients for explanatory variables $\mathrm{A}, \mathrm{B}$ and C . In addition to the main explanatory variables effects, and where the sampling design permitted, the interactions between explanatory variables were examined. The parameter $\alpha$ equates to Fulton's condition factor.

The analysis of variance tables for the linear models is presented in Table 4.3. In each case an analysis of variance was used to compare model structures. Latitude and sex were found to be non-significant effects within all models in which
they were fitted. Year effects were found to have a significant influence on model fit as both a main effects and within interaction terms with the coefficient of Log length. This indicates variation in the power relationship between length and weight within years. When age and year class effects were fitted within the model, they were found to be significant effects only if the interaction between the coefficient of Log length and the year of sampling was omitted from the model.

The final fitted model is listed in Table 4.3 as model 13. It included main effects for longitude Log length and a year effect with and interaction between year and Log length. Plots of the residual from the fitted model, model predictions and observed values are presented in Figures 4.11. Longitude has a strongly significant effect with fish collected in the Western North Sea having higher condition factors (Figure 4.12). The year effects in condition show that weight at length has been declining throughout the time period for which data is available (Fig 4.13). Recent values appear to have been more stable.

Excluding 2001, annual average bottom temperature data were available for each of the years in which the samples were taken. In order to examine correlation between the temperature signal and condition factor a model that included only Log length and the temperature time series as a linear term. The model diagnostics show that temperature has a significant effect on the fit of the model to the observations. Table 4.4. The coefficient is negative indicating a decrease in weight per unit length with increasing temperature (Figure 4.14). However, the reduction in the variance of the model fit was not as great as that achieved when fitting individual year effects to the condition factor time series. If year effects were fitted within the model the temperature effect is not significant. Table 4.4 also shows that there is a significant effect of an interaction between temperature and age on the fit of a length weight model. This suggests that there is a different effect of temperature at each age.

The results of this analysis are preliminary, no causal mechanism has been suggested and the fit is to average annual temperature. It would be expected that a temperature series recorded prior to the capture of the fish and from the location in which the fish was captured would have a dominant effect on an individual's condition.

Table 4.3. The analysis of deviance table used to fit the log weight log length generalised linear and additive models to the English groundfish survey data for North Sea cod.

| Terms | Resid. Df | RSS | Test | Df | Sum of Sq | F Value | $\operatorname{Pr}(\mathrm{F})$ | Model |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\log$ (len) | 7712 | 90.59865 |  |  |  |  |  |  |
| $\log (\operatorname{len})+$ lat | 7711 | 90.59803 | +lat | 1 | 0.000619 | 0.0527 | 0.81840 | 1 |
| $\log (\operatorname{len})+\operatorname{lon}$ | 7711 | 90.56313 | +lon | 1 | 0.035518 | 3.0242 | 0.08207 | 2 |
| $\log (\mathrm{len})+$ factor(sex) | 7710 | 90.58079 | +factor(sex) | 2 | 0.017863 | 0.7602 | 0.46759 | 3 |
| $\log (\mathrm{len})+$ factor $(\mathrm{yr})$ | 7702 | 83.75453 | +factor(yr) | 10 | 6.844125 | 62.938 | 0 | 4 |
| $\log (\mathrm{len})+$ factor(yc) | 7690 | 84.19453 | + factor(yc) | 22 | 6.404122 | 26.588 | 0 | 5 |
| $\log ($ len $)+$ factor(age) | 7708 | 90.39423 | + factor(age) | 4 | 0.204425 | 4.3579 | 0.00161 | 6 |
| $\log (\mathrm{len})+$ factor $(\mathrm{yr})$ | 7702 | 83.75453 |  |  |  |  |  |  |
| $\log (\mathrm{len}) *$ factor(yr) | 7692 | 82.21828 | $+\log (\mathrm{len}):$ factor(yr) | 10 | 1.536242 | 14.372 | 0 | 7 |
| $\log (\mathrm{len}) *$ factor(yr) | 7692 | 82.21828 |  |  |  |  |  |  |
| $\begin{aligned} & \log (\operatorname{len}) \\ & \text { factor(sex) } \end{aligned} * \quad \text { factor(yr) } \quad+$ | 7690 | 82.20466 | +factor(sex) | 2 | 0.013627 | 0.6374 | 0.5287 | 8 |
| $\log (\mathrm{len}) *$ factor $(\mathrm{yr})+$ lat | 7691 | 82.21718 | +lat | 1 | 0.001108 | 0.1037 | 0.7475 | 9 |
| $\log (\mathrm{len}) *$ factor $(\mathrm{yr})+\mathrm{lon}$ | 7691 | 82.08083 | +lon | 1 | 0.137456 | 12.880 | 0.0003 | 10 |
| $\begin{aligned} & \log (\operatorname{len}) \\ & \text { factor(age) } \end{aligned} \quad \text { factor(yr) } \quad+$ | 7688 | 82.16372 | + factor(age) | 4 | 0.054566 | 1.2764 | 0.2767 | 11 |
| $\log (\mathrm{len}) *$ factor(yr) | 7692 | 82.21828 |  |  |  |  |  |  |
| $\log (\mathrm{len}) *$ factor(yc) | 7668 | 82.72291 | 1 vs. 2 | 24 | -0.50462 | 1.9490 | 0.0036 | 12 |
| $\log (\mathrm{len}) *$ factor $(\mathrm{yr})+\mathrm{lon}$ | 7691 | 82.08083 |  |  |  |  |  |  |
| $\log (\mathrm{len}) *$ factor(yr) $+\mathrm{s}(\mathrm{lon}, 4)$ | 7687 | 81.31790 | 1 vs. 2 | 3.0 | 0.762924 | 24.037 | $1.776357 \mathrm{e}-015$ | 13 |

Table 4.4. The analysis of deviance table used to examine the effect of temperature on the fit of the log weight log length for the English groundfish survey data for North Sea cod.

| Terms |  | Resid. Df | RSS | Test | Df | Sum of Sq | F Value | $\operatorname{Pr}(\mathrm{F})$ | Model |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\log$ (len) |  | 7295 | 87.41872 |  |  |  |  |  |  |
| $\log (\operatorname{len})+$ temp |  | 7294 | 86.73512 | +temp | 1 | 0.6835991 | 57.48734 | $3.819167 \mathrm{e}-014$ | 14 |
| $\log (\operatorname{len})+$ temp |  | 7294 | 86.73512 | +temp | 1 | 0.6835991 | 57.48734 | $3.819167 \mathrm{e}-014$ |  |
| $\log$ (len) temp+factor(age):temp | + | 7290 | 86.52371 | +factor(age):temp | 4 | 0.2114185 | 4.453234 | 0.001354111 | 15 |



Figures 4.11a and 4.11b. The residual vs model prediction and observed value vs model prediction plots for a linear model of Log weight against Log length with covariates for the North Sea cod.


Figure 4.12. The response effect for the weight of North Sea cod to the position from which the English groundfish survey samples were taken. Lon - Longitude, negative values represent degrees west of the Meridian.


Figure 4.13. The response effect for the weight of North Sea cod to the year in which the English groundfish survey samples were taken.


Figure 4.14. The response effect for the weight of North Sea cod to the average annual temperature of the year in which the English groundfish survey samples were taken.

### 4.4.1 Methods and results

Data for this analysis are available both from research surveys and from market sampling, principally from the port of Lowestoft over the period 1983 to 1999. The data were extracted from the biological sampling database held at CEFAS, Lowestoft. A breakdown is given in Table 4.4. The complete dataset contains 50,000 records from market sampling and 11,654 from ground-fish surveys. Survey derived data were not considered in further analyses since the fishing gear used in the North Sea is not designed to sample plaice. For most years from 1983 onwards, the market sampling consists of between 100-300 records per month. The origin of the fish is described in (Bromley, 2000) and are principally form the central North Sea. In general individual parameters of age (from otolith analyses), length, gutted weight, sex, maturity stage have been recorded (note that plaice caught during the ground-fish survey are weighed complete). The data have been analysed by (Bromley, 2000). Plaice sexual maturation was shown to be affected by body weight, age and geographical location and was subject to annual and seasonal variation. A clear annual cycle of changes in weight at age for mature fish was demonstrated with weight increasing in the months prior to the onset of spawning (June-Dec) followed by a rapid decrease. Male fish showed a similar, though less extreme cyclical pattern. Of interest to the present working group are whether there are identifiable patterns in change of weight at length (condition) in the dataset. This was not analysed by (Bromley, 2000). Condition was computed as

Condition $=\left(\text { weight } /(\text { length })^{3}\right)^{*} 100$
Figure 4.15 shows the cycles in median condition with age averaged. A clear difference between male and female fish is apparent with the males reaching a peak in condition a month or two before the female fish.


Figure 4.15. Changes in condition of male and female fish with age averaged over all years.

Figures 4.16 and 4.17 show annual cycles in median condition for all male and female fish. There are indications of periods when peak condition appeared to be lower than average (e.g., 1991-1994 for female fish).


Figure 4.16. Annual cycles in median condition for male plaice.

## Plaice condition from market sampled data

All female fish


Figure 4.17. Annual cycles in median condition for female plaice.
Exploratory plots were made to evaluate whether there was any connection between peak female condition and peak North Sea bottom temperatures. Figure 4.18 shows that there does not appear to be any obvious relationship but it must be emphasised that this is a very preliminary and crude analysis, for example the bottom temperature data are averaged for the whole southern North Sea and not targeted from regions where plaice feed during the summer. Condition is likely affected by many other factors than temperature and this would require further exploration.

Peak female condition against maximum bottom temperature


Figure 4.18. Peak female condition against peak monthly mean bottom water temperature for whole North Sea.

Table 4.4. Summary of southern North Sea plaice data available from surveys and market sampling.

|  | From ground- <br> fish survey | From <br> market <br> sampling |
| :--- | :---: | :---: |
| 1977 Female | 404 | 0 |
| 1977 Male | 296 | 0 |
| 1978 F | 458 | 0 |
| 1978 M | 354 | 0 |
| 1979 F | 463 | 0 |
| 1979 M | 367 | 0 |
| 1980 F | 390 | 0 |
| 1980 M | 259 | 0 |
| 1981 F | 411 | 0 |
| 1981 M | 300 | 0 |
| 1982 F | 221 | 0 |
| 1982 M | 170 | 0 |
| 1983 F | 211 | 2326 |
| 1983 M | 177 | 1117 |
| 1984 F | 270 | 2285 |
| 1984 M | 207 | 1051 |
| 1985 F | 266 | 2263 |
| 1985 M | 187 | 1128 |
| 1986 F | 331 | 2221 |
| 1986 M | 252 | 993 |
| 1987 F | 276 | 2357 |
| 1987 M | 225 | 1021 |
| 1988 F | 419 | 2274 |
| 1988 M | 316 | 800 |
| 1989 F | 395 | 2379 |
| 1989 M | 314 | 841 |
| 1990 F | 464 | 1833 |
| 1990 M | 353 | 1036 |
| 1991 F | 567 | 1719 |
|  |  |  |


|  | From ground- <br> fish survey | From <br> market <br> sampling |
| :--- | :---: | :---: |
| 1991 M | 434 | 973 |
| 1992 F | 210 | 1746 |
| 1992 M | 157 | 986 |
| 1993 F | 152 | 1614 |
| 1993 M | 107 | 953 |
| 1994 F | 0 | 1484 |
| 1994 M | 0 | 902 |
| 1995 F | 0 | 1505 |
| 1995 M | 0 | 938 |
| 1996 F | 247 | 1753 |
| 1996 M | 144 | 982 |
| 1997 F | 374 | 1922 |
| 1997 M | 264 | 966 |
| 1998 F | 56 | 1851 |
| 1998 M | 53 | 1031 |
| 1999 F | 72 | 1772 |
| 1999 M | 61 | 978 |
| Total number | of | 11654 |
| observations |  | 50000 |

A preliminary exploratory analysis was undertaken to investigate the relationship between condition factor and the sexual maturity of female North Sea plaice. The maturity data were derived from the Lowestoft fish market during the period 1983-1999 and the English summer ground-fish surveys of the North Sea from 1977-1998. Data from the market sampling for months August to December ( 6,500 fish) and from the surveys for the same months ( 3,100 fish) were analysed. Market sampled fish are gutted at sea, but the gonads were left in situ. The stage of sexual maturity was ascertained for length-stratified samples of up to 200 fish (mixed sexes) per month, based on visual inspection of the exposed anterior region of the gonad. For this analysis, ovaries were classed as immature or maturing/mature. Plaice maturity was also sampled on the annual English, summer (August/September) North Sea groundfish surveys from 1977-1998, which covered the whole North Sea. The fish were caught in a Granton trawl with a 20 mm stretched-mesh liner in the cod end to retain smaller plaice than are landed commercially. Sampling was length stratified on a regional basis, with the fish being measured and weighed at sea.

Regression analysis was used to summarise the data to ascertain the general trends in maturity in relation on to year, age and condition factor. The form of the model was:

$$
\operatorname{Logit}(E[y])=I+\beta_{1} Y_{i}+\beta_{2} A_{i}+\beta_{3} C_{i}
$$

In the model the maturity of each individual fish was input in binary form, taking the value 0 if the gonad was immature and 1 if the gonad was maturing. The estimated maturity level ( $\mathrm{E}[\mathrm{y}]$ ) predicted by the model was the number of maturing and mature fish expressed as a proportion of the total number of fish, where $\mathrm{I}=$ intercept, $\mathrm{Y}=$ year, $\mathrm{A}=$ age in years and $\mathrm{C}=$ Condition factor $\left(100^{*}\right.$ Weight/Length $\left.\wedge 3\right)$. The exploratory analysis was undertaken for female fish only.

The results from market sampling (Table 4.5 and Figure 4.19) show the proportion of maturing/mature fish increased markedly with the age of the fish as expected. However, there also appears to be an effect of condition with young fish needing to reach a higher condition factor before starting to mature. Within year classes there is variability in maturity, which appears to be related to year effects and the month of sampling. It is clear that when condition factor is high, a greater proportion of fish are likely to mature, thereby increasing the spawning stock biomass estimates. Similar results are apparent in the analysis of ground-fish survey data (Table 4.6 and Figure 4.20). The shape of the relationship is slightly different probably reflecting the more restricted time over which the fish are sampled (August-September as opposed to August-December). The conclusions from the ground-fish survey data are in accord with those reached based upon market-sampled data.

Table 4.5. Logistic modelling of plaice maturity data derived from English market sampling.

## Logit(MATURITY) $=\mathbf{- 7 2 . 1 1}+\mathbf{0 . 0 3 1}$ YEAR + 1.22 AGE + 6.21 CONDITION

Analysis of Deviance

| Source | DF | Deviance | Deviance / DF | Scaled Dev | Pr $>$ |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Model | 3 | 2298.8963 | 766.2988 | 2298.8963 | $<.0001$ |
| Error | 6556 | 3243.5765 | 0.4947 | 3243.5765 |  |
| C Total | 6559 | 5542.4728 |  |  |  |

Type III (Wald) Tests
Pr >

| Source | DF | ChiSq | ChiSq |
| :--- | :--- | :--- | :--- |
| YEAR | 1 | 9.0273 | 0.0027 |
| AGE | 1 | 686.5551 | $<.0001$ |
| COND | 1 | 219.2528 | $<.0001$ |

Parameter Estimates
Pr $>$

| Variable | DF | Estimate | Std Error | ChiSq | ChiSq |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Intercept | 1 | -72.1103 | 20.7523 | 12.0743 | 0.0005 |
| YEAR | 1 | 0.0313 | 0.0104 | 9.0273 | 0.0027 |
| AGE | 1 | 1.1791 | 0.0450 | 686.5551 | $<.0001$ |
| COND | 1 | 6.2050 | 0.4191 | 219.2528 | $<.0001$ |

Table 4.6. Logistic modelling of plaice maturity data derived from English ground-fish survey.

## Logit(MATURITY) $=\mathbf{- 2 6 . 1 7}+\mathbf{0 . 0 1 0}$ YEAR + 1.04 AGE + 2.22 CONDITION

| Analysis of Deviance |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Source | DF | Deviance | Deviance / DF | Scaled Dev | Pr> |
| Model | 3 | 1164.4013 | 388.1338 | 1164.4013 | <. 0001 |
| Error | 3139 | 2839.9536 | 0.9047 | 2839.9536 |  |
| C Total | 3142 | 4004.3550 |  |  |  |
| Type III (Wald) Tests Pr > |  |  |  |  |  |
|  |  |  |  |  |  |
| Source | DF | ChiSq | ChiSq |  |  |
| YEAR | 1 | 2.0946 | 0.1478 |  |  |
| AGE | 1 | 516.5365 | <. 0001 |  |  |
| COND | 1 | 39.6058 | <. 0001 |  |  |
| Parameter Estimates$\operatorname{Pr}>$ |  |  |  |  |  |
| Variable | DF | Estimate | Std Error | ChiSq | ChiSq |
| Intercept | 1 | -26.1746 | 14.4988 | 3.2591 | 0.0710 |
| YEAR | 1 | 0.0105 | 0.0072 | 2.0946 | 0.1478 |
| AGE | 1 | 1.0457 | 0.0460 | 516.5365 | <. 0001 |
| COND | 1 | 2.2233 | 0.3533 | 39.6058 | <. 0001 |

MATUR TY predi ct ed val ues


## Condi tion fact or

Figure 4.19. English market sampling of North Sea plaice.

MATUR TY predi ct ed val ues

0.50 .60 .70 .80 .911 .11 .21 .31 .41 .51 .61 .71 .8

Condition fact or

Figure 4.20. English ground fish sampling of North Sea plaice

The allometric weight length relationship expressed as a simple condition index for plaice in the central North Sea shows clear annual cycles related to maturation and spawning. The peak of condition is reached in December for female fish and in October or November for male fish. As well as the month of sampling, condition is influenced by the sex and age of the fish as indicated by (Bromley, 2000). Statistical modelling also indicates that there is a significant year effect and there is some preliminary indication that periods of lower than average or higher than average peak condition may persist for several years. For any age group, higher condition during August - December is linked to an increased probability of ovaries maturing. The data suggest that maturity ogives used in the assessment could potentially be tuned using fish condition data. The level of condition reached in any year is presumably related to the overall feeding success of the maturing fish balanced against metabolic demands. Since temperature might be expected to affect growth and possibly levels of prey availability we undertook a crude analysis using North Sea bottom temperature data. However, this variable alone does not appear to explain inter-annual changes in peak female condition. It must be cautioned that the analysis was extremely crude and could be extended to include more direct measures of potential growth regulating factors on the feeding grounds such as prey availability or temperature data selected from the known summer feeding areas of plaice in the North Sea. Excepting temperature, the likely limit on these analyses will be the availability of time-series data for variables such as prey availability for non-coastal regions of the North Sea. The data described were derived from the English plaice fishery which takes place in the central North Sea. Similar market sampling data have also been collected from the Dutch fishery and work is currently being undertaken to jointly analyse the two data-sets (P. Bromley pers. comm.). This analysis has been complicated by differences in the maturity scales used by the two countries. Future joint studies would be greatly facilitated by the adoption of common protocols and scales for assessing maturity within ICES member countries.

### 4.4.3 Recommendations

Further work on developing the cod growth and recruitment model will focus on reconstructing the weight-at-age component. It is considered that the higher variation in observed weights in the older age classes may be partly driven by more uncertainty in their estimation, and also through additional factors influencing the weight-at-age. As such, the way that temperature influences the weight at age will be reassessed. In addition, comparing modelled and observed lengths at age will show if differences in observed and modelled weight are due to lack of fit in the growth model or varying condition factors. A possible development in the model is to allow variable condition factors, as these may be able to capture some of the year to year variation in weight at age.

The model will be used in the near future in conjunction with sea temperature projections derived from the Hadley3 coupled atmosphere-ocean model to assess possible impacts of climate change on North Sea cod over the next 50 years. This work falls within the context of longer-term scenario simulations. At present the results cannot be passed directly into medium-term projections for this stock because of the relative uncertainty in the underlying relationships included in the model.

Finally, such models could be re-parameterised for cod stocks in contrasting regions such as the Barents Sea. A comparison of stocks at the northern and southern limits of the species range could be informative.

The rationale for consideration of maternal effects (age, condition and whether fish are first-time or repeat spawners) is that these factors are very probably linked to egg quality and size and thus survivability of early life history stages. Much of the understanding of maternal factors has come from aquaculture studies. Recent mesocosm studies in Norway have attempted to extend this understanding by rearing larvae from known parental stock in mixed populations. Genetic identification was used to identify the parentage of surviving larvae. Published results from this particular study indicated that differences in environmental conditions between replicate mesocosms appeared to have a stronger effect on survival than parentage. We recommend support for further experimental studies on larval growth and survival in relation to maternal factors and also suggest that the ICES Working Group on Recruitment Processes [WGRP] might review current understanding on this subject.

Market sampling and fisheries survey observations which include measurements of condition (in whichever form is appropriate for each species) along with assessments of maturity can provide valuable insights into changes in fisheries population reproductive potential. But, such analyses must be based upon rigorously collected data. In particular for market sampled observations care must be taken to understand possible changes in selectivity of the commercial fishery. For both sources of data there needs to be unified protocols and scales for the assessment of maturity in order to ensure that data-sets from differing countries can be combined for analysis. There is a temptation to merge maturity stages to achieve this but maturity scales need to be sufficiently detailed such that patterns in maturity and spawning can still be resolved. Co-ordination of maturity assessment on national market sampling needs to be considered within ICES.

### 5.1 Medium-term projections

During the two previous meetings of SGPRISM (ICES 2000a, 2001a), work was initiated on the improvement of the prevailing ICES medium-term projection methodology, currently available as WGMTERM (Reeves and Cook, 1994) or ICP (Patterson, pers. comm.) These methods differ somewhat, but share the underlying theme of using bootstrapped recruitment residuals to drive stochastic population forecasts, with such variables as weights-at-age and proportion mature-at-age being held constant over time. The improvements proposed under the remit of SGPRISM related to the statistical appropriateness or otherwise of the bootstrap resampling approach, and the possibility of incorporating aspects of biological process modelling.

Studies have continued intersessionally, and are now pursued under the aegis of SGPRISM, the Methods Working Group (ICES 2002c) and the UK CFRD Working Group on Recruitment Processes. The development of the method is recorded in Needle et al. (2000, 2001, in press). Attention has focussed on time-series characterisations of recruitment, weights-at-age and proportion mature-at-age, because of problems caused by the way these variables are modelled in WGMTERM and its ilk. For example, the assumption of constant weights-at-age in short- and medium-term catch forecasts can result in significant estimation bias. Darby (BD4, BD5: see also ICES 2002c) has shown that for the majority of the North Sea stocks, trends in catch weights-at-age through time could have resulted in estimation bias in the total catch of up to $50 \%$. During the past five years, North Sea cod landings could have been over-estimated by 5$25 \%$, a bias that is consistent with the deficit between the total catch landed by the fishery and the final agreed TAC. If the recent landings information is "reliable", the magnitude of the bias is such that it could have removed the pressure of TAC regulation from the fishery. Darby (BD4, BD5) showed further that there appear to be systematic changes in weights-at-age that could be modelled using time series and/or cohort effect models. Removal of the bias from the short and medium-term forecasts will require modelling of such temporal changes in weight-at-age. At the current meeting of SGPRISM, two presentations were made relating to this topic (in addition to Darby, BD4 and BD5).

## Projection development work plan

The first presentation (Needle, BD6) was a summary of a paper first discussed at the recent ICES Working Group on Methods on Fish Stock Assessments [WGMG], detailing progress to date and outlining a work plan for future methodological development. Briefly, this work plan is as follows:

- Recruitment residuals will be modelled using ARMA time-series (where a significant ARMA fit is found). However, the efficacy of using only a subsample of the historical stock-recruitment pairs for each projection will be investigated. Under this scheme, several different model formulations would be fitted to these data, and the best fitting one used to generate projections. Thus, uncertainty about the form of the recruitment model would be modelled, as well as variation about it.
- The VARMA method presented in Needle (BD6) for characterising weights-at-age models incorporated cohort effects in weights, but not year effects. For stocks such as cod, cohort effects appear to be more important (Darby BD4, BD5), and in this context the VARMA model is likely to be sufficient. However, for other stocks, year effects may be more important. To account for this, an alternative has been proposed in which weights at age 1 would still be projected as residuals to a function of stock density, but subsequent ages would be modelled in terms of increments down cohorts. This would have the additional benefit of avoiding the problem in the VARMA method of ignoring data from incomplete cohorts.
- Maturity could be modelled in several ways. A single ogive could be fitted with weights as the independent variable (thus ignoring age), or use could be made of empirical nonparametric kernel distributions. Examples of these two approaches are given in Figure 5.1.1. Ultimately, deriving maturity as a function of weight will always be an approximation, and it is intended that alternative methods involving length-driven fecundity estimates will be pursued in various fora (see Section 6.1). Indeed, for many stocks, spawning-stock biomass is known to be an unsuitable proxy for egg production. The new method would base stock-recruitment modelling on potential egg production rather than spawning-stock biomass, and will thus be more cognisant of the deleterious effect of a compressed age structure (amongst other influential factors).
- Natural mortality $M$ is currently fixed throughout projections. However, annual values of $M$ are available from MSVPA and, if possible, these would be used in the new approach (after accounting for any time-series structure).
- The key justification for work on medium-term projections is to allow managers to determine the likely responses of the stock in question to specified management actions. To this end, the methodology would have the potential to allow for future imposed changes in fishing effort, gear selectivity, and catch constraints, and would enable the modelling of formal harvest control rules. WGMTERM does not currently allow for the modelling of management actions, although ICP does.
- The proposed new model will include the facility to accept a different realisation of starting values for each projection iteration, as would be produced by a bootstrapped assessment model. If such a model is not available, then starting values for each projection could be drawn from a multivariate distribution about assessment results with standard errors stipulated by the assessment model (this facility is available in both TSA and ICA).

The method would be implemented to be as general as possible, while allowing for additional model structure if data are available. It is envisaged that functional software incorporating these features, and others as need dictates, will be available by early 2003 for ICES Working Group use.

## Weights-increments modelling

The second presentation (Needle, WD8) looked at some first steps taken in the pursuance of this work plan, namely the modelling of weights as increments down cohorts. WD8 applied the approach to North Sea cod, while further work during the current SGPRISM meeting extended the analyses to North Sea haddock, whiting and plaice. Results for all these stocks are summarised below.

Three different methods of generating increments were used, namely:

1) First differences: $\quad \Delta_{a, y}=W_{a+1, y+1}-W_{a, y}$
2) Log ratios:

$$
\Delta_{a, y}=\ln \left(\frac{W_{a+1, y+1}}{W_{a, y}}\right)
$$

3) Age $N$ to age 1 ratios:

$$
\Delta_{a, y}=\frac{W_{a, 1}}{W_{a+y, y+1}}
$$

For each method, time-series along age-classes were formed (with the mean subtracted in each case) and a number of ARMA $(p, q)$ models fitted with the autoregressive order $p$ and the moving-average order $q$ both varying between 0 and 3. The significance of the fitted model parameters was evaluated using a $t$-test

$$
t=\frac{\text { parameter }}{\text { s.d.(parameter) }}
$$

and the model was accepted if $|t|>1.96$ for all fitted parameters. Where more than one model was accepted, the best was chosen on the basis of the lowest estimated value of the AIC test criterion (conditioned on the highest allowed value of $p$, in other words, 3 ).

Results of these analyses are given in Figures 5.1.2-5.1.5 and Table 5.1.1. ARMA model fits are distributed as follows:

| ARMA model order | Number of fits |  |
| :---: | :---: | :---: |
|  | 45 | $37.5 \%$ |
| 0,1 | 32 | $26.6 \%$ |
| 1,1 | 14 | $11.7 \%$ |
| 2,2 | 9 | $7.5 \%$ |
| 0,2 | 6 | $5.0 \%$ |
| 1,0 | 4 | $3.3 \%$ |
| 1,2 | 4 | $3.3 \%$ |
| 1,3 | 4 | $3.3 \%$ |
| 2,3 | 1 | $0.8 \%$ |
|  |  |  |

A sizeable percentage of weights-increments time-series do not have a significant ARMA model fit. On the other hand, there is time-series information in many of the weight-increments datasets, and in general it would seem unwise to ignore time-series structure in these data if it is present. Where no significant ARMA model could be found, the use of simple bootstrapping is probably adequate.

## North-east Arctic cod

Tretyak (WD10) presented a model of spawning-stock structure in Northeast Arctic cod, which he used to indicate which components of the spawning population have had the greatest influence on subsequent recruitment. The following is a brief summary of the method and conclusions from the Working Document.

Spawning-stock biomass (SSB) is a simple summary of the biomass of mature cod from different age groups. It takes into account only the age structure and mean weight of spawners at each age, and makes no cognisance of sex ratio, differences between the weights of males and females, absolute individual fecundity and the proportion of fish that skip spawning. It also suggests that all spawners are of the same quality and, as a result, that all eggs have a similar quality, viability, distribution and survival. WD10 proposes a hypothesis about inter- and intra-annual dissimilarity of spawners, according to which a simple summation of the biomass of mature cod from different age groups does not reveal the actual relationship between recruitment and spawners’ abundance. An algorithm for defining the abundance of firsttime and repeat spawners is presented.

The abundance of kth-time spawners at age $t$ in year $y$ can be estimated using

$$
N_{t, k, y}^{p}=N_{t, y}\left(\delta_{t-k+1, y-k+1}-\delta_{t-k, y-k}\right),
$$

where $\delta_{t, y}$ and $N_{t, y}$ are respectively the maturity-at-age and abundance-at-age for age t in year y . Coefficients of correlation between recruitment and the abundance of cod of different age spawning a different number of times indicate a weak, but stable and statistically significant positive linear relationship between recruitment and the abundance of late-maturing cod spawning for the first time at age 9 and older, as well as between recruitment and the abundance of fish at the plus-group age spawning for the 6th time or more. This permits the classification of all spawners into late-maturing (first-time spawners at age 9 and older) and early-maturing fish (all others). The first group exhibited a well-pronounced decreasing trend throughout the entire time series (Figure 5.2.1). Until the early 1960s, its abundance was close to that of the second group. Variations in these two components of the spawning stock were highly synchronous. However, at that time a change occurred in the structure of the spawning stock. The trend towards decrease in the abundance of late-maturing fish remained, while the abundance of early-maturing cod started to grow steadily.

Approximation of variations in cod recruitment by Ricker stock-recruitment models with the abundance of both spawning stock components and total spawning stock as independent variables gives the following results given in Figure 5.2.2. As expected, over the entire history of observations a statistically significant relationship exists only between recruitment and the abundance of late-maturing fish (Figure 5.2.2a). Their contribution to recruitment variance is circa $20 \%$. The contribution of early-maturing fish to the recruitment variance is below $1 \%$, which is rather counterintuitive as that component has dominated the spawning stock after the 1960s and constituted more than $93 \%$ after mid1970s (Figure 5.2.2b). This may be due to the interannual dissimilarity of early-maturing spawners which creates noise suppressing the actual dependence of recruitment, not only on the abundance of early-maturing fish, but also on the total abundance of the spawning stock (Figure 5.2.2c).

Assuming that interannual dissimilarity of the spawning stock is caused primarily by earlier maturation, we divided the time series of spawners' abundance (1949-1997) into 4 intervals. Each of these was related to a $25 \%$ interval of the logistic curve showing the age of $50 \%$ maturity (Figure 5.2.3). The 1949-1964 year-classes were assigned to the first interval, the 1965-1970 year classes to the second interval, the 1971-1975 year classes to the third interval and the 1976-1997 year classes to the fourth interval. We assumed that the age of $50 \%$ maturity approached the threshold of 6.65 years (Tretyak 2000). This led to a marked smoothing of interannual differences in quality of early-maturing spawners in the fourth interval. A decline in the abundance of older fish also reduced the dissimilarity of late-maturing cod. It can therefore be stated that interannual dissimilarity of all spawners in this interval is the lowest.

Although the abundance of late-maturing fish in the fourth interval is the lowest on the record - less than $7 \%$ of the total spawning stock - the recruitment-abundance correlation for these fish has not weakened, but even become closer (Figure 5.2.4a). For early-maturing fish this relationship acquired statistical significance when the dataset was restricted to the fourth quarter (Figure 5.2.4b). The contribution of a large amount of these fish to recruitment variance equalled that of late-maturing cod, which are only a very small fraction of the spawning stock. Addition of late-maturing to early-maturing cod does not notably change the total abundance of the spawning stock; therefore the dependence of recruitment on the total abundance became statistically significant when the dataset was restricted to the fourth quarter.

However, the total contribution of all spawners to recruitment variance did not exceed the contribution of each component of the spawning stock (Figure 5.2.4c).

The actual contribution of early- and late-maturing fish to recruitment variations is apparently somewhat greater. A determination coefficient was calculated between actual recruitment and recruitment calculated by the linear regression equation with two independent variables at the third step of recruitment approximation, and is given by

$$
\sum_{j}\left(R_{3, j}-\hat{R}_{3, j}^{(3)}\right)^{2} \Rightarrow \min
$$

where

$$
\widehat{R}_{3, j}^{(3)}=a_{1} \cdot\left(\alpha_{1} \cdot N L_{j} \cdot \exp \left(-\beta_{1} \cdot N L_{j}\right)\right)+a_{2} \cdot\left(\alpha_{2} \cdot N E_{j} \cdot \exp \left(-\beta_{2} \cdot N E_{j}\right)\right)+a_{3}
$$

Here $\mathrm{NL}_{\mathrm{j}}$ is abundance of late-maturing fish in year $\mathrm{j}, \mathrm{NE}_{\mathrm{j}}$ is abundance of early-maturing fish in year j , and $\mathrm{a}_{1}, \mathrm{a}_{2}$, and $a_{3}$ are equation parameters calculated by the integrated gradients method. For Northeast Arctic cod this coefficient was 0.563 (Figure 5.2.5). The first independent variable in the above equation is the result of the first step of approximation, that is, the approximation of recruitment by the Ricker model for late-maturing fish only. The second variable is the result of the second step, that is, approximation by the Ricker model for early-maturing fish only.

## North Sea cod and herring

Tretyak's model was applied during SGPRISM to North Sea cod and herring, in order to ascertain its general applicability and whether it would be beneficial to use it more widely.

North Sea cod abundance and recruitment estimates were taken from ICES (2002a). Maturity estimates for 1980-1995 were taken from Cook et al. (1999): however, these were rather noisy and were smoothed down cohorts using the logistic ogive $W_{a, y}=1 /(1+\exp (\alpha-\beta a))$ to prevent negative spawning abundances.

Figures 5.2.6 and 5.2.7, respectively, give the abundances of kth-time spawners for North Sea cod and herring. The relative proportions of these decline with increasing k in a straightforward manner for cod: for herring, the importance of 2 nd , 3rd, 4th and 5th time spawners fluctuates more unpredictably. This is likely to be a result of the contrasting shapes of maturity ogives for the two stocks, being quite shallow for cod and very steep for herring.

Table 5.2.1. gives linear correlation coefficients between the kth-time spawning abundance at age, and the full-stock recruitment for North Sea cod. This is not the ideal diagnostic, and should in future be augmented by (for example) an analysis of Ricker model fits to each of these relationships, but it can be seen that very young and very old spawners apparently have less influence on recruitment than those of moderate age (that is, between $\sim 5$ and $\sim 7$ years old). However, very few of these correlations are statistically significant.

North Sea herring abundance, recruitment and maturity estimates were all taken from ICES (2001e). Table 5.2.2. gives linear correlation coefficients (equivalent to Table 5.2.1) for herring. Correlation coefficients are smaller for herring than for cod (indeed, none of them are statistically significant), and there is no obvious pattern to them. This may be a consequence of the very steep maturity ogive for herring.

The Study Group concluded that Tretyak's approach to the determination of the components of the spawning population which may drive recruitment to be cogent and parsimonious, and would recommend that further analysis be conducted in the future into the model and its implications. An example of these would be the determination of the influence of first-time spawners on recruitment, with further implications for management with regards to stock rebuilding strategies.

Table 5.1.1. Orders of those best-fitting ARMA model fits to weights-increments time-series for which all fitted parameters are significant. Parameter significance was determined using a t-test, while the AIC statistic was used to select the best-fitting model. - = No significant ARMA fit.
a. North Sea cod (1963-2000).

| Age | First differences |  | Log ratios |  | Age N to age 1 ratios |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | AR | MA | AR | MA | AR | MA |
| $1-2$ | - | - | - | - | - |  |
| $2-3$ | 1 | 0 | 2 | 2 | 0 | 1 |
| $3-4$ | 0 | 3 | 2 | 3 | 0 | 3 |
| $4-5$ | - | - | 2 | 1 | 0 | 2 |
| $5-6$ | 1 | 1 | 1 | 3 | 1 | 2 |
| $6-7$ | 1 | 1 | 1 | 1 | 0 | 1 |
| $7-8$ | 1 | - | - | 1 | 1 |  |
| $8-9$ | - | 1 | - | 1 | 0 | 3 |
| $9-10$ | 1 | 1 | - | - | 1 | 3 |
| $10-11+$ | 0 |  |  |  |  | 2 |

b. North Sea haddock (1963-2000).

| Age | First differences |  | Log ratios |  | Age N to age 1 ratios |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | AR | MA | AR | MA | AR | MA |
| $0-1$ | 0 | 1 | - | - | 0 | 1 |
| $1-2$ | 0 | 2 | 0 | 2 | 0 | 1 |
| $2-3$ | 0 | 1 | 2 | 2 | 0 | 1 |
| $3-4$ | 0 | 1 | - | - | 0 | 1 |
| $4-5$ | 0 | 2 | 0 | 0 | 1 |  |
| $5-6$ | 2 | 2 | 2 | 2 | 0 | 1 |
| $6-7$ | - | - | 0 | 3 | 0 | 1 |
| $7-8$ | - | - | 1 | 0 | 0 | 1 |
| $8-9$ | - | - | - | - | 1 |  |
| $9-10+$ | - |  |  |  | 0 | 1 |

c. North Sea whiting (1960-2000).

| Ages | First differences |  | Log ratios |  | Age N to age 1 ratios |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | AR | MA | AR | MA | AR | MA |
| $1-2$ | - | - | - | - | - | - |
| $2-3$ | 2 | 2 | - | - | 0 | 1 |
| $3-4$ | - | - | - | 1 | 0 | 0 |
| $4-5$ | 1 | - | - | - | 1 |  |
| $5-6$ | - | - | - | - | - | - |
| $6-7$ | 1 | - | - | - | 0 | 3 |

d. North Sea plaice (1969-2000).

| Ages | First differences |  | Log ratios |  | Age N to age 1 ratios |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | AR | MA | AR | MA | AR | MA |
| $1-2$ | - | - | - | - | - |  |
| $2-3$ | - | - | - | 2 | 2 |  |
| $3-4$ | 0 | 1 | - | 1 | 2 | 2 |
| $4-5$ | 1 | 0 | 2 | - | - |  |
| $5-6$ | 0 | 2 | - | - | - |  |
| $6-7$ | - | - | - | - | - |  |
| $7-8$ | - | - | - | - | - |  |
| $8-9$ | 0 | 1 | - | - | - | - |
| $9-10$ | 1 | 1 | 1 | 1 | 1 |  |
| $10-11$ | 0 | 1 | 0 | - | - |  |
| $11-12$ | 0 | 2 | 1 | 1 | 1 |  |
| $12-13$ | 1 | 1 | 0 | 1 | 1 | 1 |
| $13-14$ | 0 | - | - | - | 0 | 1 |
| $14-15+$ | - |  |  |  | - | - |

Table 5.2.1. North Sea cod: linear correlation coefficients between kth-time spawner abundance at age and recruitment. Significant correlations (single-tailed test, $5 \%$ level) are highlighted in bold.

| Number of spawnings |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
|  | 1 | -0.1232 |  |  |  |  |  |  |  |  |  |  |
|  | 2 | 0.3160 | 0.0658 |  |  |  |  |  |  |  |  |  |
|  | 3 | -0.1022 | -0.3008 | -0.3086 |  |  |  |  |  |  |  |  |
|  | 4 | 0.3284 | 0.0121 | 0.2012 | 0.4149 |  |  |  |  |  |  |  |
|  | 5 | 0.3097 | 0.1814 | -0.0020 | -0.1509 | -0.0317 |  |  |  |  |  |  |
|  | 6 | 0.3066 | 0.5939 | 0.4770 | 0.3506 | 0.4445 | 0.4707 |  |  |  |  |  |
|  | 7 | 0.4247 | 0.1515 | 0.3286 | 0.3410 | 0.3791 | 0.1054 | 0.0775 |  |  |  |  |
|  | 8 | -0.4963 | -0.1123 | -0.2921 | 0.5053 | 0.7225 | 0.3393 | -0.1337 | -0.3156 |  |  |  |
|  | 9 | -0.2654 | -0.1172 | 0.5748 | 0.6154 | 0.5519 | 0.2558 | -0.0095 | -0.0134 | -0.4344 |  |  |
|  | 10 | -0.3385 | -0.3566 | -0.4070 | -0.3174 | -0.4742 | -0.1758 | -0.2009 | -0.3156 | -0.6240 | -0.4854 |  |
|  | 11 | -0.1640 | -0.1444 | -0.1665 | -0.1132 | 0.9224 | 0.5503 | 0.5161 | 0.2165 | -0.1325 | -0.2678 | -0.2476 |

Table 5.2.2. North Sea herring: linear correlation coefficients between kth-time spawner abundance at age and recruitment. There are no statistically significant correlations.

| Number of spawnings |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
|  | 1 |  |  |  |  |  |  |  |  |
|  | 2 | 0.1398 |  |  |  |  |  |  |  |
|  | 3 | -0.0395 | 0.0303 |  |  |  |  |  |  |
|  | 4 | -0.2479 | -0.1294 | -0.1099 |  |  |  |  |  |
|  | 5 |  |  | 0.1993 | 0.2035 |  |  |  |  |
|  | 6 |  |  | 0.2449 | 0.2465 | 0.2138 |  |  |  |
|  | 7 |  |  |  | -0.0634 | -0.0475 | -0.1321 |  |  |
|  | 8 |  |  |  |  | 0.1694 | -0.2870 | -0.2823 |  |
|  | 9 |  |  |  |  |  | 0.0020 | -0.1823 | -0.3103 |



Figure 5.1.1. Scatterplot of maturity-at-age against weights-at-age for North Sea cod, ages 1-6, years 1980-1995. Maturity-at-age data are taken from Cook et al. (1999), and weights-at-age data from the appropriate ICES WG report (ICES 2002a). The logistic curve is a maturity ogive fitted to all data (thus ignoring age structure). The contours are empirical kernel-smoothed distributions by age: labels indicate percentile probability.
a. Using first differences to generate weights data $\left(\Delta_{a, y}=W_{a+1, y+1}-W_{a, y}\right)$ :

b. Using log ratios to generate weights data $\left(\Delta_{a, y}=\ln \left(\frac{W_{a+1, y+1}}{W_{a, y}}\right)\right)$ :


Figure 5.1.2. Time-series of weight-at-age increments by year-classes, for North Sea cod.
c. Using ratio of age N to age 1 to generate weights data $\left(\Delta_{a, y}=\frac{W_{a, 1}}{W_{a+y, y+1}}\right)$ :


Figure 5.1.2 Continued. Time-series of weight-at-age increments by year-classes, for North Sea cod.
a. Using first differences to generate weights data $\left(\Delta_{a, y}=W_{a+1, y+1}-W_{a, y}\right)$ :

b. Using log ratios to generate weights data $\left(\Delta_{a, y}=\ln \left(\frac{W_{a+1, y+1}}{W_{a, y}}\right)\right)$ :


Figure 5.1.3. Time-series of weight-at-age increments by year-classes, for North Sea haddock.
c. Using ratio of age N to age 1 to generate weights data $\left(\Delta_{a, y}=\frac{W_{a, 1}}{W_{a+y, y+1}}\right)$ :





Age $\mathrm{a}=7$





Figure 5.1.3 Continued. Time-series of weight-at-age increments by year-classes, for North Sea haddock.
a. Using first differences to generate weights data $\left(\Delta_{a, y}=W_{a+1, y+1}-W_{a, y}\right)$ :

b. Using log ratios to generate weights data $\left(\Delta_{a, y}=\ln \left(\frac{W_{a+1, y+1}}{W_{a, y}}\right)\right)$ :








Figure 5.1.4. Time-series of weight-at-age increments by year-classes, for North Sea whiting.
c. Using ratio of age N to age 1 to generate weights data $\left(\Delta_{a, y}=\frac{W_{a, 1}}{W_{a+y, y+1}}\right)$ :


Figure 5.1.4 Continued. Time-series of weight-at-age increments by year-classes, for North Sea whiting.
a. Using first differences to generate weights data $\left(\Delta_{a, y}=W_{a+1, y+1}-W_{a, y}\right)$ :

b. Using log ratios to generate weights data $\left(\Delta_{a, y}=\ln \left(\frac{W_{a+1, y+1}}{W_{a, y}}\right)\right)$ :


Figure 5.1.5. Time-series of weight-at-age increments by year-classes, for North Sea plaice.
c. Using ratio of age N to age 1 to generate weights data $\left(\Delta_{a, y}=\frac{W_{a, 1}}{W_{a+y, y+1}}\right)$ :


Figure 5.1.5 Continued. Time-series of weight-at-age increments by year-classes, for North Sea plaice.


Figure 5.2.1. Northeast Arctic cod.


Figure 5.2.2 Ricker stock-recruitment model fits for Northeast Arctic cod (1949-1997), using the abundance of a) late maturing and b) early maturing fish, as well as c) total abundance.


Figure 5.2.3. Age of $\mathbf{5 0 \%}$ maturity ( $\mathbf{t}_{\mathbf{s}}$ ) in different year classes of cod

Figure 5.2.3. Northeast Arctic cod.


Figure 5.2.4. Ricker stock-recruitment model fits for Northeast Arctic cod (1976-1997), using the abundance of a) late maturing and b) early maturing fish, as well as c) total abundance.


Figure 5.2.5. Sequential approximation of recruitment to Northeast Arctic cod by linear regression equations with two independent variables.


Figures 5.2.6. Abundance of North Sea cod spawning for the kth time. The legend gives values of k .


Figure 5.2.7. Abundance of North Sea herring spawning for the kth time. The legend gives values of $k$.

## RECOMMENDATIONS

### 6.1 Recommendations for future work

This third meeting of SGPRISM is to be its last. The proposed work plan for development of medium-term projection software and methodology (see Section 5.1) calls for such software to be finalised by early 2003, for subsequent ICES certification and use by stock assessment Working Groups during that year. Before this programming exercise can be completed, there needs to be a thorough collation and evaluation of the appropriate data available for each stock in the ICES area, along with analyses of which process models would be suitable for incorporation in stock projections. It is probable that this work could only realistically be undertaken by a dedicated Study Group - the alternative would be for stock assessment Working Groups to collate data and formulate process models, but to demand this of them would be unreasonable given their current workloads. To this end, SGPRISM makes the following recommendation.

A Study Group on Growth, Maturity and Condition Indices in Stock Projections [SGGROMAT] (Co-chairs: C. L. Needle, UK and C. T. Marshall, Norway) be established to meet at ICES Headquarters, Copenhagen, Denmark from 510 December 2002 to:
a) collate data on weights, maturity, condition, fecundity, and age-length and length-weight keys for stocks in the North Sea, Irish Sea, Northeast Arctic and Baltic Sea;
b) develop the implementation of growth, maturity and condition models for use in projections for those stocks for which data are available; and
c) agree an intersessional programme to apply the findings of the Study Group to areas not covered by ToR a).

SGGROMAT will report for the attention of the Resource Management, Living Resources, Oceanography and Baltic Committees. The report will be brought to the attention of the Working Group on Methods on Fish Stock Assessments [WGMG] and ACFM.

Intersessional work implementing growth and condition models for North Sea and Northeast Arctic cod is planned for the summer of 2002 (as a collaborative effort between scientists from Aberdeen and Bergen), and it is intended that SGGROMAT will follow this lead and apply it to other stocks where possible. The key justification for SGGROMAT is that, without it, assessment Working Groups will not be able to use the new medium-term projection methodology. If this is so, they will need to revert to such approaches as WGMTERM, and Darby (BD4, BD5) has shown that assuming fixed weights-at-age in particular can seriously degrade stock projections. The output from SGGROMAT would take the form of Working Papers to each assessment Working Group, proposing candidate projection models for application to stocks within their remit.

Participants at SGGROMAT should include process modellers, assessment scientists, and data collators. The purpose of locating the meeting at ICES Headquarters in Copenhagen, Denmark is to permit ready access to all available assessment Working Group reports and datasets, along with such relevant sources as the IBTS database. SGGROMAT should precede the January 2003 meeting of WGMG, which will review it and recommend (or otherwise) its findings to assessment Working Groups meeting in 2003.

SGPRISM would recommend further that the Working Group on Recruitment Processes [WGRP] pursues process studies in such a way that outcomes and results can be used to drive projection models.

### 6.2 Links to other groups and activities

In this section is a partial listing of on-going international collaborative groups investigating aspects of the reproductive variability of marine fish stocks, biological processes contributing to recruitment and numerical approaches to quantifying uncertainty in medium-term stock forecasts.

### 6.2.1 NAFO Working Group on Reproductive Potential

Methods for quantifying the reproductive potential of individuals and stocks are currently being reviewed and summarized by the NAFO Working Group on Reproductive Potential (Chair: E.A. Trippel, Canada). A third Working Group meeting has been proposed for autumn 2002. It is anticipated that the information compiled by the working group will benefit fisheries management by developing methods for incorporating relevant information on stock structure into conventional assessment.

### 6.2.2

The ICES Working Group on Recruitment Processes [WGRP] (Co-chairs: P. Pepin, Canada, and R. Nash, Isle of Man) worked by correspondence in 2001 but will meet in 2002 to address the following tasks.
a) Review multidisciplinary projects dealing with recruitment research, with attention to providing a synthesis of the projects and highlight unresolved issues, which deserve further consideration.
b) Consider the results of the SGPRISM's examination of the STEREO project, along with concurrent and subsequent investigations.
c) Evaluate an analysis of simulations exploring the effects of stock structural factors on the parameters of stockrecruitment relationships with a view to preparing a case for a Study Group on evaluating the impact of these factors on stock projections, and in the light of results from the Workshop on the Transport of Cod Larvae held immediately prior to the meeting.
d) Consider a synthesis of the 2000 Theme Session on Spatial and Temporal Patterns in Recruitment Processes to be prepared by the Session's Convenors as well as a synthesis of recruitment issues presented at the SAP symposium held in December 2000.
e) Identify areas in the study of recruitment processes where sufficient progress has been achieved with the objective of developing a workshop dealing with specific case studies.
f) Review the development of new approaches or techniques used in the study of factors and processes that influence the development and survival of fish eggs and larvae in relation to recruitment or the formation of year-class strength.

The report from the second meeting of SGPRISM (ICES 2001a) contains an examination of the STEREO project which will allow the Working Group on Recruitment Processes to address its second ToR b). In addition, this report of the third meeting of SGPRISM includes a number of concurrent and subsequent investigations, which should lead to a refinement of the present approach to incorporating stock structure and environmental information into stock assessment and projection procedures.

## SGPRISM suggests that drivers for weight-at-age and maturity-at-age should be investigated on a stock by stock

 basis.
### 6.2.3 ICES International Bottom Trawl Survey Working Group

The ICES International Bottom Trawl Survey Working Group [IBTSWG] (Chair: A.W. Newton, UK) meets in 2002 with a ToR j) to consider the additional collection of data on the condition of cod (liver weights) caught during the first quarter IBTS in the North Sea and recommend a protocol on how to collect such data. The collection of such data, which could be used to evaluate the reproductive potential of cod stocks, was a recommendation from the second meeting of SGPRISM (ICES 2001a).

Field and laboratory studies of fish growth indicate that condition varies rapidly within and across years in response to changes in the balance between energy intake (e.g., feeding) and expenditure (e.g., spawning). This has important implications for the reproductive potential of demersal stocks such as the Barents Sea and Icelandic cod where the fecundity/length relationships are significantly affected by inter-annual variation in the condition of spawners. For many stocks it is difficult to assess the magnitude of variability in condition because condition time series are non-existent or of limited temporal/spatial scope. In such cases, the sampling protocols of research surveys should be adapted such that individual-level information on total body and/or liver weights is collected on an annual basis for selected species. When adapting the sampling protocols special consideration should be given to the following:
a) how the condition index should be formulated for the species of interest (e.g., weight-based versus lipid-based); and
b) how sub-sampling strategies as well as the temporal and spatial coverage of the survey affect the accuracy of the condition index.

### 6.2.4 ICES Working Group on Methods on Fish Stock Assessments

The work of SGPRISM has relevance to the ICES Working Group on Methods on Fish Stock Assessments; specifically in respect of models for weight-at-age, maturity-at-age and condition factors for both historical stock assessment and medium-term projections.

### 6.2.5 ICES Study Group on the Further Development of the Precautionary Approach to Fishery Management

The ICES Study Group on the Further Development of the Precautionary Approach to Fishery Management [SGPA] (Co-chairs: Colin Bannister, UK, and Manuela Azevedo, Portugal) meets in 2002 with a ToR:
b) to continue the development of the framework for formulating advice for:
i) stocks under full analytical assessment

- where the reference points are based on $\mathrm{F}_{\text {loss }}$ and $\mathrm{B}_{\text {loss }}$
- where the reference points are based on historical evidence of reduced recruitment at low SSB levels

The segmented regression approach presented at this meeting by O'Brien and Maxwell (WD3, WD4) potentially provides an objective technique whereby biomass reference points might be developed. SGPRISM proposes that the approach be investigated further with a view to addressing the ToR b) of the SGPA.

### 6.2.6 ICES Planning Group on North Sea Cod and Plaice Egg Surveys

The ICES Planning Group on North Sea Cod and Plaice Egg Surveys [PGEGGS] (Chair: John Casey, UK) meets in 2002 with a ToR b) to co-ordinate the timing, area allocation and sampling protocol for plankton surveys; including sampling of adults to determine fecundity, atresia, maturity and sex ratio. SGPRISM suggests that this planning group be made aware of the current SGPRISM report, particularly in respect of the Section 2 which discusses reproductive potential for North Sea cod.

## 7 WORKING DOCUMENTS AND BACKGROUND MATERIAL PRESENTED TO THE STUDY GROUP

At the Study Group twelve working documents (WD1-WD12) and six background documents (BD1-BD6) were presented and discussed. These are listed below, together with the reference codes used in the text of this report.

Working document: WD1

Bogstad, B. Predation by herring on capelin larvae - incorporation in capelin management.

Working document: WD2

Witthames, P.R. Reproduction and stock evaluation for recovery (RASER).

Working document: WD3
O'Brien, C.M., and Maxwell, D.L. Stock-recruitment modelling based upon a segmented regression approach - the case of North Sea cod.

Working document: WD4

O'Brien, C.M., and Maxwell, D.L. Stock-recruitment modelling based upon a segmented regression approach - the case of Bay of Biscay anchovy.

Working document: WD5
Bogstad, B. Incorporating process information in recruitment predictions for Northeast Arctic cod.

Working document: WD6

Nash, R.D.M., and Dickey-Collas, M. At what point in the life history is year class strength determined? An examination of North Sea herring (Clupea harengus).

Working document: WD7

Armstrong, M., Dickey-Collas, M., and Gerritsen, H. Estimation of reproductive characteristics of Irish Sea cod.

Working document: WD8

Needle, C. Exploratory analyses: weights-at-age modelling
Working document: WD9
Marshall, C.T., Yaragina, N.A., Thorsen, A., and Kjesbu, O.S. Incorporating reproductive potential into stockrecruitment models for Northeast Arctic cod.

Working document: WD10

Tretyak, V.L. Spawning stock structure of Northeast Arctic cod.
Working document: WD11
Clark, R., and Fox, C. Modelling the effects of climatic change on North Sea cod stocks.
Working document: WD12
Reeves, S.A. Possible approaches to using process information in medium-term projections for Eastern Baltic cod.
Background document: BD1
Report of the $2^{\text {nd }}$ Working Group Meeting, NAFO Scientific Council Working Group on Reproductive Potential, St. Petersburg, Russia, 23-26 October 2001.

Background document: BD2

NAFO Working Group on Reproductive Potential - current and proposed terms of reference.
Background document: BD3

NAFO Working Group on Reproductive Potential - data on reproductive potential for Atlantic cod in ICES sub-areas I and II.

Background document: BD4
Darby, C. Estimation bias in the North Sea short-term forecasts. Working paper presented to the ICES Working Group on Methods on Fish Stock Assessments, ICES Headquarters, Copenhagen, Denmark, 3-7 December 2001.

Background document: BD5
Darby, C. Over-estimation bias in the North Sea cod short-term forecasts. Working paper presented to the ICES Working Group on Methods on Fish Stock Assessments, ICES Headquarters, Copenhagen, Denmark, 3-7 December 2001.

Background document: BD6

Needle, C.L. ARMA and VARMA models in medium-term projections. Working paper presented to the ICES Working Group on Methods on Fish Stock Assessments, ICES Headquarters, Copenhagen, Denmark, 3-7 December 2001.

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APPENDIX A: AVAILABLE DATA ON REPRODUCTIVE POTENTIAL OF NORTHEAST ARCTIC COD. ADAPTED FROM TABLES PREPARED BY THE NAFO WORKING GROUP ON REPRODUCTIVE POTENTIAL

| Common |  |
| :--- | :--- |
| name: | ATLANTIC COD |

Area: $\square$ Stock:
NORTHEAST ARCTIC

Created by:
C.T. MARSHALL/N.A. YARAGINA
2001-08

## Species:

Gadus morhua
name:

ICES SUB-AREAS I AND II

Updated by:


| Data status |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Stock size | Stock composition | Age | Sex <br> ratio | Maturity | Fecundity | Weight | Condition | $\begin{gathered} \text { Additional } \\ \text { data } \end{gathered}$ |
| 2001 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ |  |
| 2000 | $\sqrt{ }$ | $\checkmark$ | $\checkmark$ | $(\sqrt{ })$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\checkmark$ |  |
| 1999 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |  |
| 1998 | $\sqrt{ }$ | $\checkmark$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ |  |
| 1997 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ |  |
| 1996 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ |  |
| 1995 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ |  |
| 1994 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ |  |
| 1993 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1992 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1991 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ }$ ) | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1990 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ }$ ) | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1989 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1988 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1987 | $\sqrt{ }$ | $\checkmark$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1986 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1985 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1984 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1983 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1982 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1981 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1980 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1979 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1978 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1977 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1976 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1975 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1974 | $\checkmark$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\checkmark$ | $\sqrt{ }$ |
| 1973 | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $(\sqrt{ })$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |


| Data status |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Stock size | Stock composition | Age | $\begin{array}{\|c} \text { Sex } \\ \text { ratio } \\ \hline \end{array}$ | Maturity | Fecundity | Weight | Condition | $\begin{gathered} \text { Additional } \\ \text { data } \end{gathered}$ |
| 1972 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 1971 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 1970 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 1969 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 1968 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 1967 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 1966 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 1965 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 1964 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 1963 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 1962 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\sqrt{ }$ | $\checkmark$ | $\checkmark$ |
| 1961 | $\checkmark$ | $\checkmark$ | $\checkmark$ | ( V ) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 1960 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 1959 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 1958 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |
| 1957 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |
| 1956 | $\checkmark$ | $\checkmark$ | $\checkmark$ | ( V ) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |
| 1955 | $\checkmark$ | $\checkmark$ | $\checkmark$ | ( $\sqrt{ }$ ) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |
| 1954 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |
| 1953 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |
| 1952 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |
| 1951 | $\checkmark$ | $\checkmark$ | $\checkmark$ | ( V ) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |
| 1950 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |
| 1949 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |
| 1948 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |
| 1947 | $\checkmark$ | $\checkmark$ | $\checkmark$ | (V) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |
| 1946 | $\checkmark$ | $\checkmark$ | $\checkmark$ | ( V ) | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |
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APPENDIX B: BASIS, QUALITY AND FORMAT OF DATA ON REPRODUCTIVE POTENTIAL OF NORTHEAST ARCTIC COD. ADAPTED FROM TABLES PREPARED BY THE NAFO WORKING GROUP ON REPRODUCTIVE POTENTIAL


## RECOMMENDED METHOD FOR FECUNDITY ANALYSIS:

GRAVIMETRIC, STEREOMETRIC OR AUTO-DIAMETRIC

| Ref. no.: | 1 |
| :--- | :--- |
| Ref. no.: | 2 |
| Ref. no.: | 3 |

Ref. no.: $\quad 4,5$

| Data basis, format and quality |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Variables | Year range | Data basis (A/L/W) | Data origin | Sampling frequency | Notes on data, methods and contents | Ref. No. |
| Stock size | 1946-2001 | A | CC, S | Q,A | VPA | 6 |
| Stock composition | $\begin{aligned} & 1978-2001 \\ & 1985-2001 \\ & 1946-2000 \\ & 1932-1979 \end{aligned}$ | $\begin{gathered} \text { L,A,W } \\ \text { L,A,W } \\ \text { L,A,W } \\ \text { L,A } \end{gathered}$ | CC, S | Q | Winter survey <br> Lofoten survey <br> Russian survey <br> Norwegian sampling of commercial fishery | $\begin{array}{\|l\|} \hline 6 \\ 6 \\ 6,7 \\ 8 \end{array}$ |
| Age determination | 1946-2001 | A | CC, S | Q | Otoliths | 33, 50 |
| Sex ratio | $\begin{aligned} & \text { 1985-1996 } \\ & \text { 1950's- } \\ & \text { 1990's } \\ & \hline \end{aligned}$ | L <br> A | S $\mathrm{CC}, \mathrm{~S}$ | A <br> 10 YEAR MEANS | Winter + Lofoten surveys combined |  |
| Maturity: |  |  |  |  |  |  |
| A. Ogives (E) | 1946-2001 | A | S | A |  | 6,10 |
| B. Skip of spawning | $\begin{array}{\|l\|} \hline 1985-1996 \\ 1991-1992 \\ \hline \end{array}$ | $\begin{gathered} \mathrm{L} \\ \mathrm{~L}, \mathrm{~A} \\ \hline \end{gathered}$ | $\begin{array}{\|l} \hline \mathrm{S} \\ \mathrm{~S} \\ \hline \end{array}$ | $\begin{aligned} & \hline \mathrm{A} \\ & \mathrm{~A} \\ & \hline \end{aligned}$ |  | $\begin{array}{\|l\|} \hline 9 \\ 51 \\ \hline \end{array}$ |
| C. Spawning probability |  |  |  |  |  |  |
| D. Other | $\begin{aligned} & 1923-1976 \\ & \text { 1989-1997 } \\ & \text { 1950's- } \\ & 1990 \text { 's } \\ & \text { 1940's- } \\ & 1950 \text { 's } \\ & \hline \end{aligned}$ | $\begin{gathered} \mathrm{A}, \mathrm{~L} \\ \mathrm{~A}, \mathrm{~L} \\ \mathrm{~A} \end{gathered}$ | $\begin{aligned} & \mathrm{CC} \\ & \mathrm{~S} \\ & \mathrm{CC}, \mathrm{~S} \end{aligned}$ | M | Spawning check in otolith sex differences in ogive sex differences in ogive <br> Sorokin maturity scale | $\begin{aligned} & \hline 11 \\ & 12 \\ & 10 \\ & 56,57 \end{aligned}$ |


| Data basis, format and quality |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Variables | Year range | Data basis <br> (A/L/W) | Data origin | $\underset{\substack{\text { Sampling } \\ \text { frequency }}}{ }$ | Notes on data, methods and contents | Ref. No. |
| A. Estimation | $\begin{aligned} & \hline 1986-1991 \\ & 1999 \\ & 1971-1972 \\ & 1990, \\ & 1991, \\ & 1993- \\ & 1996,1999 \\ & \hline \end{aligned}$ | $\begin{gathered} \hline \text { L, W } \\ \text { L, W } \\ \text { L,W,A } \\ \text { L,W,A } \end{gathered}$ | $\begin{aligned} & \text { EW } \\ & \text { EW } \end{aligned}$ |  | Gravimetric auto-diametric gravimetric gravimetric | $\begin{aligned} & \hline 4 \\ & 5 \\ & 42 \\ & 52 \\ & \hline \end{aligned}$ |
| B. First time vs. repeat spawners | 1989-1992 | L,W | X |  | Spawners held in captivity consecutive seasons over | 13,14 |
| C. Atresia | lab expt | L,W | EC |  | Laboratory-reared cod | 15 |
| D. Other | lab expt |  | EC |  | Effects of starvation on relative fecundity | 20 |
| Weight: |  |  |  |  |  |  |
| A. Commercial fisheries data | $\begin{array}{\|l\|} \hline 1946-2001 \\ 1949-1993 \\ \hline \end{array}$ | $\begin{aligned} & \hline \mathrm{A} \\ & \mathrm{~A} \\ & \hline \end{aligned}$ | $\begin{aligned} & \hline \mathrm{CC}, \mathrm{~S} \\ & \mathrm{CC}, \mathrm{~S} \end{aligned}$ | $\begin{aligned} & \mathrm{A} \\ & \mathrm{~A} \\ & \hline \end{aligned}$ | Russian+Norwegian Russian | $\begin{array}{\|l} \hline 6,8 \\ 53 \\ \hline \end{array}$ |
| B. Survey data | 1984-2001 | A | S | A | Russian+Norwegian | 6 |
| C. Other | 1913-1953 | L | CL | A | Historical data on size from Lofoten | 47 |
| Condition: |  |  |  |  |  |  |
| A. Fulton |  |  |  |  |  |  |
| B. HSI | $\begin{array}{\|l\|} \hline 1927-1996 \\ 1828-1842 \\ \hline \end{array}$ | L | $\begin{aligned} & \mathrm{CC} \\ & \mathrm{CL} \end{aligned}$ | $\begin{gathered} \mathrm{M} \\ \mathrm{~A} \\ \hline \end{gathered}$ | By $10-\mathrm{cm}$ length class industrial index of liver | $\begin{array}{\|l\|} \hline 16 \\ 48 \\ \hline \end{array}$ |
| C. Energy | 1946-1996 | L | CC, S | A | Total lipid energy in livers of mature females | 17 |
| D. Other | 1929-1982 | W | CL | A | median intensity samples | 2 |
| Egg viability: |  |  |  |  |  |  |
| A. Egg quality | lab expt <br> lab expt <br> lab expt <br> lab expt <br> lab expt <br> lab expt |  | $\begin{aligned} & \mathrm{EC} \\ & \mathrm{EC} \\ & \mathrm{EC} \\ & \mathrm{EC} \\ & \mathrm{EC} \\ & \mathrm{EW} \end{aligned}$ |  | Cortical reaction following fertilization egg diameter and dry weight <br> egg diameter in relation to maternal size <br> egg diameter in relation to maternal condition <br> egg dry weight in relation to spawning experience buoyancy studies | 18 <br> 19 <br> 1 <br> 15 <br> 13 <br> 30 |
| B. Fertilisation success | lab expt <br> lab expt <br> lab expt |  | EC <br> EC <br> EW |  | Incomplete hardening of chorion associated with poor fertilization links <br> between broodstock, egg quality and fertilization <br> links between length and condition of females and fertilization | 18 <br> 21 $27$ |


| Data basis, format and quality |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Variables | Year range | Data basis $(\mathrm{A} / \mathrm{L} / \mathrm{W})$ | Data origin | Sampling frequency | Notes on data, methods and contents | Ref. No. |
| C. Egg mortality | $\begin{aligned} & \text { lab expt } \\ & 1983-1984 \end{aligned}$ |  | EC <br> S | A | Links between egg mortality and spawning experience <br> field estimate of egg mortality | $14$ $22$ |
| D. Other | review <br> review <br> review <br> lab expt |  | EC |  | Broodstock management and egg rearing egg quality maternal effects fatty acid composition | $\begin{aligned} & \hline 23 \\ & 24 \\ & 25 \\ & 26 \\ & \hline \end{aligned}$ |
| Larval viability: |  |  |  |  |  |  |
| A. Hatching success |  |  |  |  |  |  |
| B. Larvae quality | mesocosm <br> mesocosm |  | EC <br> EC |  | Relationship between larval growth and survival and links to broodstock <br> relationship between $\underset{\text { survival }}{\text { larval }}$ growth and | $28$ $29$ |
| C. Mortality | $\begin{array}{\|l\|} \hline 1983-1985 \\ 1983-1985 \end{array}$ |  | S |  | Mortality and firstfeeding mortality estimates from ichthyoplankton surveys | $\begin{aligned} & 31 \\ & 32 \end{aligned}$ |
| D. Other |  |  |  |  |  |  |
| Spawning time | consult original reference |  |  |  | ref. 34 describes seasonal activity of thyroid gland | 2,3,34 |
| Contamination | 1993 |  | S | A | Concentration of heavy metals, PAH, and organochlorine discussed in relation to egg abnormalities | 35 |
| Environmental key factors | consult original reference |  |  |  | Temperature, wind stress | $\begin{aligned} & 36, \\ & 37,54 \end{aligned}$ |
| Other factors or parameters | consult original reference |  |  |  | Cannibalism, overwintering of recently settled juveniles historical data on Barents Sea | $\begin{array}{\|l} 38 \\ 39 \\ 55 \end{array}$ |


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| :--- | :--- |
|  |  |
|  |  |
| Area: |  |
| Stock: |  |
|  |  |

## Estimation of reproductive potential

| Subject | Brief description | Year range | Ref. No. |
| :--- | :--- | :--- | :--- |
| Potential or realised <br> egg production | Estimates of total egg production (also referred to as <br> population fecundity) | $1985-1996$ | $9,40,41$, <br> 42, <br> 44 |
| Viable egg and <br> larvae production | Estimates of relative abundance of eggs and larvae from <br> Russian ichthyoplankton surveys conducted from April-July | $1959-1993$ | 45 |
| Critical life stages | Early pelagic juveniles and 0-group surveys |  | 32,46 |
| Environmental <br> influences | Influence of wind and temperature on S/R relationship | 37 |  |
| Stock recruitment <br> relations | Biomass-based S/R relationship <br> bioenergetic index of reproductive potential | $1946-1989$ <br> $1946-1996$ | 37 |
| Other studies | Reconstructing the S/R relationship using the total lipid <br> energy contained in the livers of mature females | $1946-1996$ | 17,49 |

## APPENDIX D: DATA SOURCES FOR NORTHEAST ARCTIC COD. ADAPTED FROM TABLES PREPARED BY THE NAFO WORKING GROUP ON REPRODUCTIVE POTENTIAL

| $\xrightarrow[\mathrm{N}]{\mathbf{C}}$ | ATLANTIC COD |  |
| :---: | :---: | :---: |
| Area: | ICES SUB-AREAS I AND II |  |
| Stock: | NORTHEAST ARCTIC |  |
| Stock size | 6 | 6. ICES 2001. Report of the Arctic Fisheries Working Group. ICES CM 2001/ACFM:19. |
| Stock composition | 6,7,8 | 7. Ponomarenko, V.P., Ponomarenko, I.Ya., and Yaragina, N.A. 1980. Growth and maturation of the Lofoten-Barents Sea cod. ICES CM 1980/G:25. <br> 8. Jørgensen, T. 1992. Long-term changes in growth of North-east Arctic cod (Gadus morhua) and some environmental influences. ICES J. mar. Sci. 49: 263-277. |
| Age determination | 33 50 | 33. Rollefsen, G. 1933. The otoliths of the cod. Fiskeridirektoratets skrifter, serie havunders $\varnothing$ kelser 4: 1-14. <br> 50. Mankevich E.M., 1966. Methods of taking and reading the age samples of cod. Materially rybocoz. issledov. Severnogo basseina, Murmansk, v.7, p.53-56. (In Russian). |
| Sex ratio | 9 | 9. Marshall, C.T., Kjesbu, O.S., Yaragina, N.A., Solemdal, P., and Ulltang, Ø. 1998. Is spawner biomass a sensitive measure of the reproductive and recruitment potential of Northeast Arctic cod? Can. J. Fish. |
| Maturity: |  |  |
| A. Ogives (E) | 10 | 10. Ponomarenko, I.Ya., and Yaragina, N.A. 1994. Maturity rate of the Lofoten-Barents Sea cod in 40's and 90's. ICES CM 1994/G:30. |
| B. Skip of spawning | 9, 51 | 51. Oganesyan S.A., 1993. Periodicity of the Barents Sea cod reproduction. ICES C.M. 1993/G:64 |
| C. Spawning probability |  |  |
| D. Other |  | 11. Jørgensen, T. Long-term changes in age at sexual maturity of Northeast Arctic cod (Gadus morhua L.). J. Cons. int. Explor. Mer 46: 235-248. <br> 12. Ajiad, A., Jakobsen, T., and Nakken, O. 1999. Sexual difference in maturation of Northeast Arctic cod. J. Northw. Atl. Fish. Sci. 25: 1-15. <br> 56. Sorokin V.P., 1957. Ovogenesis and sexual cycle in cod (Gadus morhua morhua L.) Trudy PINRO, 10: $125-144$ (in Russian). (English translation No.72F49, Ministry of Agriculture, Fisheries and Food, United Kingdom 1961). <br> 57. Sorokin V.P., 1960. Sexual cycle and spermatogenesis in cod. Trudy PINRO, 12: 71- 87 (in Russian). |


| Fecundity: |  |  |
| :---: | :---: | :---: |
| A. Estimation | 4,5,42,52 | 4. Kjesbu, O.S., Witthames, P.R., Solemdal, P., and Greer Walker, M. 1998. Temporal variations in the fecundity of Arcto-Norwegian cod (Gadus morhua) in response to natural changes in food and temperature. J. Sea Res. 40: 303-321. <br> 5. Thorsen, A., and Kjesbu, O.S. in press. A rapid method for estimation of follicle size and potential fecundity in Atlantic cod using a computeraided particle analysis system. J. Sea Res. <br> 52. Kiseleva V.M., 2000. Individual and Population fecundity of the Barents Sea cod in the 90 's. Materials of the session of PINRO on the results of research in 1998-1999. Murmansk: PINRO Press, Part 1, p.10-22. (in Russian). |
| B. First time vs. repeat spawners | 13,14 | 13. Kjesbu, O.S., Solemdal, P., Bratland, P., and Fonn, M. 1996. Variation in annual egg production in individual captive Atlantic cod (Gadus morhua). Can. J. Fish. Aquat. Sci. 53:610-620. <br> 14. Solemdal, P., Kjesbu, O.S., and Fonn, M. 1995. Egg mortality in recruit- and repeat-spawning cod - an experimental study. ICES CM 1995/G:35. |
| C. Atresia | 15 | 15. Kjesbu, O.S., Klungsøyr, J., Kryvi, H., Witthames, P.R., and Greer Walker, M. 1991. Fecundity, atresia, and egg size of captive Atlantic cod (Gadus morhua) in relation to proximate body composition. Can. J. Fish. Aquat. Sci. 48: 2333-2343. |
| D. Other | 20 | 20. Karlsen, Ø., Holm, J.C., and Kjesbu, O.S. 1995. Effects of periodic starvation on reproductive investment in first-time spawning Atlantic cod (Gadus morhua L.). Aquaculture 133: 159-170. |
| Weight: |  |  |
| A. Commercial fisheries data | 53 | 53. Ozhigin V.K., Tretyak V.L., Yaragina N.A., Ivshin V.A., 1994. Dependence of the Barents Sea cod growth upon conditions of their feeding on capelin and water temperature. ICES C.M. 1994/G:32 |
| B. Survey data |  |  |
| C. Other | 47 | 47. Rollefesen, G. 1954. Observations on the cod and cod fisheries of Lofoten. Rapp. P.-v. Réun. Cons. perm. int. Explor. mer 136: 40-47. |
| Condition: |  |  |
| A. Fulton |  |  |
| B. HSI | 16, 48 | 16. Yaragina, N.A., and Marshall, C.T. 2000. Trophic influences on interannual and seasonal variation in the liver condition index of Northeast Arctic cod (Gadus morhua). ICES J. Mar. Sci. 57:42-55. <br> 48. Øiestad, V. 1994. Historic changes in cod stocks and cod fisheries: Northeast Arctic cod. ICES mar. Sci. Symp. 198: 17-30. |
| C. Energy | 17 | 17. Marshall, C.T., Yaragina, N.A., Ådlandsvik, B., and Dolgov, A.V. 2000. Reconstructing the stock-recruit relationship for Northeast Arctic cod using a bioenergetic index of reproductive potential. Can. J. Fish. Aquat. Sci. 57:2433-2442. |
| D. Other |  |  |


| Egg viability: |  |  |
| :---: | :---: | :---: |
| A. Egg quality | $\begin{array}{ll} 18, & 19 \\ 30 \end{array}$ | 18. Kjørsvik, E., and Lønning, S. 1983. Effects of egg quality on normal fertilization and early development of cod, Gadus morhua L. J. Fish Biol. 23: 1-12. <br> 19. Knutsen, G.M. and Tilseth, S. 1985. Growth, development, and feeding success of Atlantic cod larvae Gadus morhua related to egg size. Trans. Am. Fish. Soc. 114: 507-511. <br> 30. Davenport, J., Lønning, S., and Kjørsvik, E. 1981. Osmotic and structural changes during early development of eggs and larvae of the cod, Gadus morhua L. J. Fish Biol. 19: 317-331. |
| B. Fertilisation success | 21, 27 | 21. Kjørsvik, E., Stene, A., and Lønning, S. 1984. Morphological, physiological and genetical studies of egg quality in cod (Gadus morhua L.) In: E. Dahl, D.S. Danielssen, E. Moksness and P. Solemdal (ed.), The propagation of cod Gadus morhua L. Flødevigen rapportser. 1: 67-86. <br> 27. Solemdal, P. and others. 1993. Size of spawning Arcto-Norwegian cod (Gadus morhua L.) and the effects on their eggs and early larvae. ICES CM 1993/G:41. |
| C. Egg mortality | 22 | 22. Fossum, P. 1988. A tentative method to estimate mortality in the egg and early fish larval stages with special reference to cod (Gadus morhua L.) FiskDir. Skr. Ser. HavUnders. 18: 329-349. |
| D. Other | $\begin{aligned} & 23,24, \\ & 25 \end{aligned}$ | 23. Kjørsvik, E., and Holmefjord, I. 1995. Atlantic halibut (Hippoglossus hippoglossus) and cod (Gadus morhua). In: N.R. Bromage and R.J. Roberts (ed.), Broodstock management and egg and larval quality. Blackwell Science, Cambridge. p 169-196. <br> 24. Kjørsvik, E., Mango-Jensen, A. and Holmefjord, I. 1990. Egg quality in fishes. Adv. Mar. Biol. 26: 71-113. <br> 25. Solemdal, P. Maternal effects - a link between the past and the future. J. Sea Res. 37: 213-227. <br> 26. Ulvund, K.A., and Grahl-Nielsen, O. 1988. Fatty acid composition in eggs of Atlantic cod (Gadus morhua). Can. J. Fish. Aquat. Sci. 45: 898-901. |
| Larval viability: |  |  |
| A. Hatching success |  |  |
| B. Larvae quality | 28, 29 | 28. Blom, G., Svåsand, T., Jørstad, K.E., Otterå, H., Paulsen, O.I., and Holm, J.C. 1994. Comparative survival and growth of two strains of Atlantic cod (Gadus morhua) through the early life stages in a marine pond. Can. J. Fish. Aquat. Sci. 51: 1012-1023. <br> 29. van der Meeren, T., Jørstad, K.E., Solemdal, P., and Kjesbu, O.S. 1994. Growth and survival of cod larvae (Gadus morhua L.): comparative enclosure studies of Northeast Arctic cod and coastal cod from western Norway. ICES mar. Sci. Symp. 198: 633-645. |
| C. Mortality | 31, 32 | 31. Ellertsen, B., Fossum, P., Solemdal, P., and Sundby, S. 1989. Relation between temperature and survival of eggs and first-feeding larvae of northeast Arctic cod (Gadus morhua L.). Rapp. P.-v- Réun. Cons. int. Explor. Mer 1991: 209-219. <br> 32. Sundby, S., Bjørke, H., Soldal, A. V., and Olsen, S. 1989. Mortality rates during the early life stages and year-class strength of northeast Arctic cod (Gadus morhua L.) Rapp. P.-v. Réun. Cons. int. Explor. Mer 191: 351-358. |
| D. Other |  |  |


| Spawning time | 2, 34 | 2. Pedersen, T. 1984. Variation in peak spawning of Arcto-Norwegian cod (Gadus morhua L.) during the time period 1929-1982 based on indices estimated from fishery statistics. In: E. Dahl, D.S. Danielssen, E. Moksness and P. Solemdal (editors), The propagation of cod Gadus morhua L. Flødevigen rapportser. 1: 301-316. <br> 34. Woodhead, A.D. 1959. Variations in the activity of the thyroid gland of the cod, Gadus callarias L., in relation to its migrations in the Barents Sea. J. mar. biol. Ass. U.K. 38: 407-415. |
| :---: | :---: | :---: |
| Contamination | 35 | 35. Mukhina, N.V., Plotitsyna, N.F., and Golubeva, T.A. 1996. Disturbances in embryogenesis of cod from the Lofoten-Barents Sea stock. ICES C.M./Q:6. |
| Environmental key factors | $\begin{array}{ll} \hline 36, & 37, \\ 38, & 39, \\ 54 & \\ \hline \end{array}$ | 36. Sundby, S. 2000. Recruitment of Atlantic cod stocks in relation to temperature and advection of copepod populations. Sarsia 85: 277-298. <br> 37. Ottersen, G. and Sundby, S. 1995. Effects of temperature, wind and spawning stock biomass on recruitment of Arcto-Norwegian cod. Fish. Oceanogr. 4: 278-292. <br> 38. Bogstad, B. and Mehl, S. 1997. Interactions between Atlantic cod (Gadus morhua) and its prey species in the Barents Sea. In Proceedings of the international symposium on the role of forage fishes in marine ecosystems. Alaska Sea Grant College Program Report No. 97-01, pp 591-615. <br> 39. Ponomarenko, I.Ya. 1984. Survival of bottom-dwelling young cod in the Barents Sea and its determining factors. In: O.R. Godø and S. Tilseth (editors), Reproduction and recruitment of Arctic cod. Proceedings of the Soviet-Norwegian symposium, Leningrad 26-30 September 1983. pp. 213-229. <br> 54. Tretyak V.L., Ozhigin V.K., Yaragina N.A., Ivshin V.A., 1995. Role of oceanographic conditions in Arcto-Norwegian cod recruitment dynamics. ICES C.M. 1995/ Mini:15. Mini-symposium on Arctic Oceanographic Processes. |


| Estimation of reproductive potential: |  |
| :--- | :--- | :--- |
| A. Potential or realised egg <br> production | 40. Sundby, S. and Solemdal, P. 1984. The egg production of Arcot- <br> Norwegian cod (Gadus morhua L.) in the Lofoten area estimated by egg <br> surveys. In: O.R. Godø and S. Tilseth (editors), Reproduction and <br> recruitment of Arctic cod. Proceedings of the Soviet-Norwegian <br> symposium, Leningrad 26-30 September 1983. pp. 116-137. <br> 41. Sundby, S. and Bratland, P. 1987. Spatial distribution and <br> production of eggs from Northeast-arctic cod at the coast of northern <br> Norway 1983-1985. Fisken Hav. 1987: 1-58 (in Norwegian). <br> 42. Aldonov, V.K., Borisov, V.M., and Serebryakov, V.P. 1982. <br> Fecundity of Arcto-Norwegian cod spawning populations. ICES <br> CM/G:19. <br> 43. Serebryakov, V.P. 1990. Population fecundity and reproductive <br> capacity of some food fishes in relation to year-class strength <br> fluctuations. J. Cons. int. Explor. Mer 47: 267-272. <br> 44. Serebryakov, V.P. 1991. Predicting year-class strength under <br> uncertainties related to survival in the early life history of some North <br> Atlantic commercial fish. NAFO Sci. Counc. Studies 16: 49-55. |
| B. Viable egg and <br> larvae production | 45. Mukhina, N.V. 1999. The use of Russian ichthyoplankton survey <br> data in the forecasting of recruitment to Arcto-Norwegian cod stock. <br> ICES CM 1999/Y:15. |


| C. Critical life stages | 32,46 | 46. Helle, K., Bogstad, B., Marshall, C.T., Michalsen, K., Ottersen, G., <br> and Pennington, M. 2000. An evaluation of recruitment indices for <br> Arcto-Norwegian cod (Gadus morhua L.) Fish. Res. 48: 55-67. |
| :--- | :--- | :--- |
| D. Environmental <br> influences |  |  |
| E. Stock recruitment <br> relations |  | Other references <br> 1,3, 49, <br> 55 |
| 1. Kjesbu, O.S. 1989. The spawning activity of cod, Gadus morhua L. J. <br> Fish Biol. 34: 195-206. <br> and ecology of the gadoid resources of the Barents Sea. Fish. Res. 5: <br> 119-161. <br> 49. Nilssen, E.M., Pedersen, T., Hopkins, C.C.E., Thyholt, K., and <br> Pope, J.G. 1994. Recruitment variability and growth of Northeast Arctic <br> cod: influence of physical environment, demography, and predator-prey <br> energetics. ICES mar. Sci. Symp. 198: 449-470. <br> 55. Maslov N.A., 1944. The bottom-fishes of the Barents Sea and their <br> fisheries. Transactions of PINRO, vol. 8, p. 3-186 (In Russian). |  |  |

APPENDIX E: AVAILABLE DATA ON REPRODUCTIVE POTENTIAL OF NORTH SEA COD. ADAPTED FROM TABLES PREPARED BY THE NAFO WORKING GROUP ON REPRODUCTIVE POTENTIAL


| Year | Stock size | Stock composition | Age | Sex ratio | Maturity | Fecundity | Weight | Condition | Additional data |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 |  |  |  |  |  |  |  |  |  |
| 2000 | $\sqrt{ }$ | $\sqrt{ }$ |  | $\sqrt{ }$ |  |  |  |  |  |
| 1999 | $\checkmark$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |
| 1998 | $\checkmark$ | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |  |  |  |
| 1997 | $\sqrt{ }$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ |  |  |  |  |
| 1996 | $\sqrt{ }$ | $\checkmark$ |  | $\sqrt{ }$ | $\sqrt{ }$ |  |  |  |  |
| 1995 | $\sqrt{ }$ | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |  |  |  |
| 1994 | $\checkmark$ | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |  |  |  |
| 1993 | $\checkmark$ | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |  |  |  |
| 1992 | $\sqrt{ }$ | $\sqrt{ }$ |  | $\checkmark$ | $\checkmark$ |  |  |  |  |
| 1991 | $\sqrt{ }$ | $\checkmark$ |  | $\checkmark$ | $\sqrt{ }$ |  |  |  |  |
| 1990 | $\sqrt{ }$ | $\sqrt{ }$ |  | $\sqrt{ }$ | $\sqrt{ }$ |  | $\sqrt{ }$ |  |  |
| 1989 | $\checkmark$ | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  | $\checkmark$ |  |  |
| 1988 | $\sqrt{ }$ | $\checkmark$ | $\sqrt{ }$ | $\sqrt{ }$ | $\sqrt{ }$ |  | $\checkmark$ |  |  |
| 1987 | $\sqrt{ }$ |  | $\checkmark$ | $\checkmark$ | $\checkmark$ |  |  |  |  |
| 1986 | $\sqrt{ }$ | $\sqrt{ }$ |  | $\checkmark$ | $\sqrt{ }$ |  |  |  |  |
| 1985 | $\checkmark$ | $\checkmark$ |  | $\checkmark$ | $\sqrt{ }$ |  |  |  |  |
| 1984 | $\checkmark$ | $\checkmark$ |  | $\checkmark$ | $\sqrt{ }$ |  |  |  |  |
| 1983 | $\sqrt{ }$ | $\checkmark$ |  | $\checkmark$ | $\checkmark$ |  |  |  |  |
| 1982 | $\checkmark$ | $\checkmark$ |  | $\checkmark$ | $\sqrt{ }$ |  |  |  |  |
| 1981 | $\checkmark$ | $\checkmark$ |  | $\checkmark$ | $\sqrt{ }$ |  |  |  |  |
| 1980 | $\sqrt{ }$ | $\checkmark$ |  | $\sqrt{ }$ | $\sqrt{ }$ |  |  |  |  |
| 1979 | $\checkmark$ | $\checkmark$ |  |  |  |  |  |  |  |
| 1978 | $\sqrt{ }$ | $\sqrt{ }$ |  |  |  |  |  |  |  |
| 1977 | $\sqrt{ }$ | $\checkmark$ |  |  |  |  |  |  |  |
| 1976 | $\sqrt{ }$ | $\checkmark$ |  |  |  |  |  |  |  |
| 1975 | $\sqrt{ }$ | $\checkmark$ |  |  |  |  |  |  |  |
| 1974 | $\sqrt{ }$ | $\sqrt{ }$ |  |  |  |  |  |  |  |
| 1973 | $\sqrt{ }$ | $\sqrt{ }$ |  |  |  |  |  |  |  |
| 1972 | $\sqrt{ }$ | $\sqrt{ }$ |  |  | $\sqrt{ }$ | $\sqrt{ }$ |  |  |  |
| 1971 | $\checkmark$ | $\checkmark$ | $\sqrt{ }$ |  |  | $\checkmark$ |  |  |  |


| Year | Stock | $\begin{gathered} \text { Stock } \\ \text { composition } \end{gathered}$ | Age | $\underset{\text { ratio }}{\text { Sex }}$ | Maturity | Fecundity | Weight | Condition | Additional |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | $\checkmark$ | $\checkmark$ | - |  |  | $\checkmark$ | $\checkmark$ |  |  |
| 1969 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | $\checkmark$ | $\checkmark$ | $\checkmark$ | - |  |
| 1968 | $\checkmark$ | , | $\checkmark$ | - | $\checkmark$ | - | - | - | - |
| 1967 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - |  | - | - |  |  |
| 1966 | $\checkmark$ | $\checkmark$ | $\checkmark$ |  | - | - | , | - |  |
| 1965 | $\checkmark$ | - | $\checkmark$ |  | - | - | - | - |  |
| 1964 | $\checkmark$ | $\checkmark$ |  |  |  | - | - |  |  |
| 1963 | $\checkmark$ | $\checkmark$ |  |  |  |  |  |  |  |
| 1962 |  |  |  |  |  |  |  |  |  |
| 1961 |  |  |  |  |  |  |  |  |  |
| 1960 |  |  |  |  |  |  |  |  |  |
| 1934 | - |  | - | - | $\checkmark$ |  | - | - |  |
| 1893 | - | - | - | - | $\checkmark$ |  | - | - |  |
| 1891 | - | - | - | - |  | $\checkmark$ | - |  |  |
| 1920-63 | $\checkmark$ | $\checkmark$ |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
|  | - | - | - | E |  | - | - |  |  |
|  | - | $\square$ | - |  | - |  | - | - |  |
|  |  |  |  |  |  |  | - |  |  |
|  | - |  | - | - |  |  | - |  |  |
|  | - | - | - | - |  |  | - |  |  |
|  |  | - |  |  |  |  | - | - |  |
|  |  |  |  |  |  |  |  |  |  |
|  | - |  | - | $\square$ | - | - | - |  |  |
|  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  | - |  |  |  |
|  | - |  | - | E |  |  | - |  |  |
|  |  |  | - |  |  |  | - |  |  |
|  | - |  |  |  |  |  | - |  |  |
|  |  |  | - |  |  |  |  | - |  |
|  |  |  |  |  |  |  |  |  |  |
|  |  |  | - |  |  | - |  | - |  |
|  | - |  | - | - |  |  | - |  |  |
|  | - | - | - | - | - | - | - | - | - |

APPENDIX F: BASIS, QUALITY AND FORMAT OF DATA ON REPRODUCTIVE POTENTIAL OF NORTH SEA COD. ADAPTED FROM TABLES PREPARED BY THE NAFO WORKING GROUP ON REPRODUCTIVE POTENTIAL

| STOCK: |  |
| :--- | :--- |
|  | NORTHEAST ARCTIC |
| AREA: | ICES AREA IV |

REPRODUCTIVE STRATEGY: DETERMINATE BATCH SPAWNER

Timing of spawning:

| JANUARY -APRIL |
| :---: |
| DECEMBER-JANUARY |

Ref. no.: 16

DECEMBER-JANUARY

## RECOMMENDED METHOD FOR FECUNDITY ANALYSIS:

Pre-Breeding season and estimate seasonal atresia
Ref. no.: 6,16

## Data basis, format and quality

| Parameters | Year range | Data basis (A/L/W) | Data origin | Sampling frequency | Notes on data, methods and contents | Ref. No. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stock size | $\begin{array}{\|l\|} \hline 1920-1963 \\ 1963-2000 \end{array}$ | $\begin{aligned} & \mathrm{A} \\ & \hline \mathrm{~A} \end{aligned}$ | $\mathrm{CL}, \mathrm{CC}$ in CL, some areas after 1990 | $\begin{aligned} & \mathrm{M} \\ & \mathrm{M} \end{aligned}$ | VPA but high uncertainty in estimates Catch at age analysis (VPA) | $\begin{aligned} & 20 \\ & 1 \end{aligned}$ |
| Stock composition | 1963 2000 1970 s, 198890, 1999 | $\begin{gathered} \mathrm{L}, \mathrm{~A} \\ \mathrm{~L}, \mathrm{~A}, \mathrm{~W} \end{gathered}$ | $\begin{aligned} & \mathrm{CL} \\ & \mathrm{CL}, \mathrm{CC} \end{aligned}$ | M | Catch at age analysis (VPA) <br> No routine weight measurements measurements | $\begin{array}{ll} \hline 1 \\ 11, \quad 12, \\ 13 \end{array}$ |
| Age determination | $\begin{array}{ll} 1963 & - \\ 2000 & - \end{array}$ | otoliths | C,L | M | age based assessments | 1 |
| Sex ratio | 1963-2000 | L, A | S | A | Surveys | 13 |
| Maturity: |  |  |  |  |  |  |
| A. Ogives | 1883, 1924, $1968-72$ $1980-1995$ $1995-2001$ | A,L, W | S | A | dedicated study dedicated study dedicated study analysis from ICES IBTS analysis of survey data in progress P. Wright, FRS | $\begin{aligned} & \hline 2 \\ & 3 \\ & 4 \\ & 5 \\ & 5 \\ & 13 \end{aligned}$ |
| $\begin{array}{\|lr\|} \hline \begin{array}{l} \text { B. } \\ \text { spawning } \end{array} & \text { of } \\ \hline \end{array}$ |  |  |  |  |  |  |
| C. Spawning probability | 1953-90 |  | S | A | from egg surveys review | $6$ |
| Fecundity: |  |  |  |  |  |  |
| A. Estimation | $\begin{array}{\|l} \hline<1891 \\ 1969 \\ 1970-72 \\ 1999 \\ \hline \end{array}$ | $\begin{gathered} \mathrm{A}, \mathrm{~L} \\ \mathrm{~A}, \mathrm{~L} \\ \mathrm{~A}, \mathrm{~L}, \mathrm{~W} \\ \mathrm{~A}, \mathrm{~L}, \mathrm{~W} \\ \hline \end{gathered}$ | $\begin{array}{\|l\|l\|} \hline \mathrm{S} \\ \mathrm{~S} \\ \hline \end{array}$ | $\begin{aligned} & \text { A } \\ & \text { A } \\ & \hline \end{aligned}$ | $\begin{aligned} & \mathrm{n}=92 \\ & \mathrm{n}=26 \end{aligned}$ | $\begin{array}{\|l\|} \hline 8 \\ 9 \\ 4 \\ 10 \\ \hline \end{array}$ |
| B. First time vs. repeat spawners |  |  |  |  |  |  |
| C. Atresia |  |  |  |  |  |  |
| D. Other |  |  |  |  |  |  |

## Data basis, format and quality

| Parameters | Year range | Data basis (A/L/W) | Data origin | Sampling frequency | Notes on data, methods and contents | Ref. No. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Weight: |  |  |  |  |  |  |
| A. Commercial fisheries data | $\begin{array}{\|l} \hline 1970 \mathrm{~s} \\ 1988-90 \end{array}$ |  | CC |  | no regular sampling | $\begin{aligned} & 11 \\ & 11 \\ & 12 \end{aligned}$ |
| B. Survey data |  |  | S |  |  |  |
| C. Other |  |  |  |  |  |  |
| Condition: |  |  |  |  |  |  |
| A. Fulton |  |  |  |  |  |  |
| B. HSI | $\begin{array}{\|l} \hline 1969-70 \\ 1999 \\ \hline \end{array}$ |  | CC, S | M | few data $1^{\text {st }}$ quarter | $\begin{array}{\|l\|} \hline 11 \\ 10 \\ \hline \end{array}$ |
| C. Energy |  |  |  |  |  |  |
| D. Other |  |  |  |  |  |  |
| Egg viability: |  |  |  |  |  |  |
| A. Egg quality | 1970-72 | A,L,W | S | A | $\mathrm{n}=30$ weights | 4 |
| B. Fertilisation <br> success      |  |  |  |  |  |  |
| C. Egg mortality |  |  |  |  | size selective mortality | 14 |
| D. Other |  |  |  |  |  |  |
| Larval viability: |  |  |  |  |  |  |
| A. Hatchingsuccess |  |  |  |  |  |  |
| B.Buality Larvae |  |  |  |  |  |  |
| C. Mortality |  |  |  |  |  |  |
| D. Other |  |  |  |  | $\begin{aligned} & \text { study } \\ & \text { related } \\ & \text { development }\end{aligned} \quad \begin{aligned} & \text { temperature } \\ & \text { larval }\end{aligned}$ development | 15 |
| Spawning time | 1953-90 |  | S | A | plankton surveys | 5, 16 |
| Contamination |  |  |  |  |  |  |
| Environmental key factors |  |  | S |  | regional trophic studies temperature effects on recruitment | $\begin{aligned} & 17,18 \\ & 19 \end{aligned}$ |
| Other factors or parameters |  |  |  |  |  |  |

APPENDIX G: STUDIES OF REPRODUCTIVE POTENTIAL OF NORTH SEA COD. ADAPTED FROM TABLES PREPARED BY THE NAFO WORKING GROUP ON REPRODUCTIVE POTENTIAL.


| Estimation of reproductive potential |  |  |  |
| :--- | :--- | :--- | :--- |
| Subject | Short description | Year range | Ref. No. |
| Potential or realised <br> egg production |  |  |  |
| Viable egg and <br> larvae production |  | $1962-1993$ | 21 |
| Critical life stages | REVIEW OF GADOID STOCK CHANGES |  |  |
| Environmental <br> influences |  | $1962-1993$ | 21 |
| Stock recruitment <br> relations | REVIEW OF GADOID STOCK CHANGES |  |  |
| Other studies |  |  |  |

## APPENDIX H: DATA SOURCES FOR NORTH SEA COD. ADAPTED FROM TABLES PREPARED BY THE NAFO WORKING GROUP ON REPRODUCTIVE POTENTIAL.



| Data sources |  |  |
| :---: | :---: | :---: |
| Parameters | Ref. No. | Data sources <br> (literature reference or contact person) |
| Stock size | 20 | ICES (2001) Report of the ICES Advisory Committee on Fishery Management, 2000. ICES Co-operative Research Report No 242, 361-366 Pope, J. G., and Macer, C. T. (1996) An evaluation of the stock structure of North Sea cod, haddock and whiting since 1920, together with a consideration of the impact of fisheries and predation effects on their biomass and recruitment, 53, 1157-1169. |
| Stock composition | 1 | ICES (2001) Report of the ICES Advisory Committee on Fishery Management, 2000. ICES Cooperative Research Report No 242, 361-366. |
| Age determination | 1 | ICES (2001) Report of the ICES Advisory Committee on Fishery Management, 2000. ICES Cooperative Research Report No 242, 361-366. |
| Sex ratio |  |  |
| Maturity: |  |  |
| A. Ogives | 2 3 4 5 | Holt, E.W.L (1893) North Sea Investigations. Journal of the Marine Biological association, UK, 78-122 <br> Graham, M. (1924) The annual cycle of the mature cod in the North Sea Fishery Investigations, London (2) 6,1-77. <br> Oosthuizen and Daan (1974) Egg fecundity and maturity of North Sea cod, Gadus morhua. Netherlands Journal of Sea Research 8(4), 378-397 <br> Cook, R.M., Kunzlik, P.A., Hislop, J.R.G. and Poulding, D. (1999) Models of growth and maturity for North Sea cod, Journal of the Northwest Atlantic Fisheries Science 25, 91-99. |
| B. Skip of spawning |  |  |
| C. Spawning probability |  | Brander, K.M. (1994) The location and timing of cod spawning around the British Isles <br> ICES (1994). Spawning and life history information for North Atlantic cod stocks, ed. K. Brander. ICES Cooperative Research Report 205 |
| D. Other |  |  |
| Fecundity: |  |  |
| A. Estimation | 4 10 | Fulton, T.W. (1891). The comparative fecundity of sea fishes. Report of the Fishery Board of Scotland, 9, 243-268. <br> Schopka, S.A. (1971), Fortpflanzungrate bei Herings- und Kabeljaupopulationen. Ber dt. Wiss. Kommn. f Meeresforsch N.F. 22, 31-79 <br> Oosthuizen and Daan (1974) Egg fecundity and maturity of North Sea cod, Gadus morhua. Netherlands Journal of Sea Research 8(4), 378-397 <br> Wright, P. unpublished data |
| $\begin{array}{\|l\|} \hline \text { B. First time vs. } \\ \text { repeat spawners } \\ \hline \text { C. Atresia } \end{array}$ |  |  |
|  |  |  |


| Data sources |  |  |
| :---: | :---: | :---: |
| Parameters | Ref. No. | Data sources <br> (literature reference or contact person) |
| D. Other |  |  |
| Weight: |  |  |
| A. Commercial fisheries data | $\begin{aligned} & 11 \\ & 12 \end{aligned}$ | FRS Marine Laboratory, Aberdeen, Scotland CEFAS Fishery laboratory, Lowestoft, England other EU institutes |
| B. Survey data | 13 | ICES International bottom trawl surveys, H. Sparholt, ICES Headquarters |
| C. Other |  |  |
| Condition: |  |  |
| A. Fulton | 11, 12, 1 | available from individual length and weight data sources |
| B. HIS |  | liver weights 1969,70 and 1999- FRS Marine Laboratory, Aberdeen, Scotland |
| C. Energy |  |  |
| D. Other |  |  |
| Egg viability: |  |  |
| A. Egg quality | 4 | Oosthuizen and Daan (1974) Egg fecundity and maturity of North Sea cod, Gadus morhua. Netherlands Journal of Sea Research 8(4), 378-397 |
| B. <br> success Fertilisation |  |  |
| C. Egg mortality | 14 | Rijnsdorp, A. D., and Jaworski (1990) A. Size-selective mortality in plaice and cod eggs: a new method in the study of egg mortality. Journal du Conseil International pour l'Exploration de la mer 47, 256-263. |
| D. Other |  |  |
| Larval viability: |  |  |
| A. Hatching success |  |  |
| B. Larvae quality |  |  |
| C. Mortality |  |  |
| D. Other | 15 | Thompson, B.M., and Riley, J.D. (1981). Egg and larval development studies in the North Sea cod (Gadus morhua L.) Rapp. P-v. Reun. Cons. Int Explor. Mer 178, 553-559. |
| Spawning time | $16$ | Brander, K.M. (1994) The location and timing of cod spawning around the British Isles <br> Hislop, J.R.G. (1984) A comparison of the reproductive tactics and strategies of cod, haddock whiting and Norway pout in the North Sea. |
| Contamination |  |  |
| Environmental key factors | 17 18 19 | Munk. P. (1997). Prey size spectra and prey availability of larval and small juvenile cod. J. Fish. Biol. 51(Suppl.A): 340-351. <br> Munk. P., Larsson, P.O., Danielssen, D., and Moksness, E. (1995). Larval and small juvenile cod Gadus morhua concentrated in the highly productive areas of a shelf break front. Mar. Ecol. Prog. Ser. 125: 21-30. <br> Planque, B. \& Frédou, T (1999). Temperature and the recruitment of Atlantic cod (Gadus morhua). Canadian Journal of Fisheries and Aquatic Science 56, 2069-2077. |
| Estimation of reproductive potential: |  |  |
| A. Potential or <br> realised egg <br> production  |  |  |
| B. Viable egg and larvae production |  |  |


| Data sources |  | Ref. No. |
| :--- | :--- | :--- |
| Data sources <br> (literature reference or contact person) |  |  |
| C. Critical life <br> stages | 21 | ICES (2001) Workshop on Gadoid stocks in the North sea during the <br> 1960s and 1970s. The Fourth ICES/GLOBEC Backward Facing Workshop <br> ICES Cooperative Research Report 244, pp55. |
| D. Environmental <br> influences |  |  |
| E. Stock <br> recruitment <br> relations |  |  |
| Other references |  |  |

