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International Council for the
C.M.1992/Assess:15 Exploration of the Sea

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## REPORT OF THE WORKING GROUP ON NORTH ATLANTIC SALMON

Dublin, Ireland, 5-12 March 1992

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Run Reconstruction Modelling

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## 1 INTRODUCTION

### 1.1 Main Tasks

At its 1991 Statutory Meeting, ICES resolved (C.Res. 1991/2:7:3) that the Working Group on North Atlantic Salmon (Chairman: Dr. K. Friedland) should meet in Dublin, Ireland from 5-12 March 1992 to consider questions which include those posed to ICES by NASCO (Appendix 1).

Two Study Groups and one Workshop met prior to the Working Group and submitted reports: The Study Group on the Norwegian Sea and Faroes Salmon Fishery, the Study Group on the North American Salmon Fisheries, and the Workshop on Salmon Assessment Methodology.

The Working Group considered a further 23 papers submitted by participants (Appendix 2). References cited in the report are given in Appendix 3.

### 1.2 Participants

| Baum, E.T. | USA |
| :--- | :--- |
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| Crozier, W.W. | UK (N. Ireland) |
| Dunkley, D.A. | UK (Scotland) |
| Friedland, K. (Chairman) | USA |
| Hansen, L.P. | Norway |
| Holm, M. | Norway |
| Ikonen, E. | Finland |
| Karlsson, L. | Sweden |
| Marshall, T.L. | Canada |
| Meerburg, D.J. | Canada |
| Møller Jensen, J. | Denmark |
| O'Maoileidigh, N. | Ireland |
| Porter, T.R. | Canada |
| Potter, E.C.E. | UK (England \& Wales) |
| Prévost, E. | France |
| Rago, P. | USA |
| Reddin, D.G. | Canada |
| Sharov, A. | Russia |
| Zubchenko, A. | Russia |

## 2 <br> CATCHES OF NORTH ATLANTIC SALMON

### 2.1 Nominal Catches of Salmon

Total nominal catches of salmon reported by country in all fisheries for 1960-1991 are given in Table 2.1.1, and nominal catches in homewater fisheries for 1960-1991 are given in Table 2.1.2.

Catch statistics in the North Atlantic area also include fish farm escapees and in the North-East area ranched fish. The updated total catch for 1990 of $4,890 t$ is
$1,003 \mathrm{t}$ less than the total catch in 1989 of $5,893 \mathrm{t}$. Total landings for 1990 were the lowest recorded and show decreases for several countries. Figures for 1991 ( $4,031 \mathrm{t}$ ) are provisional, but it appears likely that the final data will still show a decrease from 1990. This is the fifth year in which the total catch has decreased from the previous year. The decline in the catch of wild stocks may be greater than suggested by the catch statistics because of the inclusion in the statistics of increasing catches of fish farm escapees and ranched fish.

The lack of information on fishing effort presents major difficulties in interpreting the catch data of any one year and also in comparing catches of different years. Management plans in several countries are designed to decrease catches. The trends in catch data will be discussed in Section 3.

### 2.2 Catches in Numbers by Sea Age and Weight

Reported nominal salmon catches for several countries by sea age and weight are summarized in Table 2.2.1. As in Tables 2.1.1 and 2.1.2, catches in some countries include both wild and reared salmon and fish farm escapees. Figures for 1991 are provisional. The methods used by the different countries to break down their total catch by sea age are described in Anon. (1986 and 1987).

### 2.3 Unreported Catches

### 2.3.1 Unreported catches within commission areas

Unreported catches by year and commission area, as estimated by the Working Group, are presented in Table 2.3.1 except for the West Greenland Commission Area which are unavailable. The total unreported catch in 1991 was estimated to be $1,682 \mathrm{t}$; a decrease of $29 \%$ from the five year mean of $2,377 \mathrm{t}$ (Table 2.3.1). Unreported catch estimated for the North-East Commission area was $1,555 \mathrm{t}$ in 1991; a decrease of $30 \%$ from the five year mean, 1987-91. Unreported catches estimated for the North American Commission was 127 t in 1991; a decrease of $32 \%$ from the six year mean of 187 t . The estimates of non-reported catches in previous Working Group Reports were provisional and have been revised in Table 2.3.1. Consequently, it is impossible to derive estimates for the West Greenland Commission by subtraction of values reported in this report from those previously reported values. However, total non-catch fishing mortalities, which include unreported catches, have been estimated for the West Greenland Commission area (range of values 0.1 to 0.3 ).

### 2.3.2 Unreported catches in international waters

The 1989/90 estimate of unreported catch in international waters in the North East Atlantic Commission area has
been updated to reflect new information. The unreported catch is estimated to be between 180 t and 350 t .

Activity in this area was greatly reduced in 1990/91 with only one or two vessels thought to have been operating. The catch in this area may have been between 25 and 100 t (Anon. 1992b). There are no known catches of salmon in international waters in either the North American Commission Area (Anon. 1992a) or the West Greenland Commission Area.

## 3 STATUS OF STOCKS OCCURRING IN COMMISSION AREAS

### 3.1 Organization of Stock Status Information

At its 1991 meeting, the Working Group (Anon 1991a) recommended that participants bring long-term datasets that could be considered as indicators of stock status and sustainability. The information that follows consists of data that have been presented at previous Working Group meetings, with additional data that were tabled at the 1992 meetings of the Study Group on North American Salmon Fisheries and the Study Group on the Norwegian Sea and Faroes Salmon Fishery. The Working Group attempted to present information on stock status in a more comprehensive fashion and to depict trends where they occur.

### 3.2 Eastern North Atlantic

### 3.2.1 Measures of abundance

## Catches

The total nominal landings of salmon in the north east Atlantic during the period 1960 to 1991, including the European fraction of the Greenland catch, are provided in Figure 3.2.1.1. The landings increased from more than $5,000 \mathrm{t}$ in 1960 , peaked at nearly $9,000 \mathrm{t}$ in the beginning of the 1970s, and decreased towards 1991 when the landings were about $3,500 \mathrm{t}$, the lowest during the period.

Trends in total nominal catches by weight (cf. Table 2.1.1) were analyzed by country from 1960 to 1991 , and during the period 1976 to 1990 using linear regression (Table 3.2.1.1). In the first period, the catches decreased significantly by year in Norway, Russia, N. Ireland and Scotland, whereas there was a significant ( $\mathrm{P}<0.05$ ) increase in catches in Iceland. In Finland, France, Ireland and Sweden, no trends were apparent. During the last 15 years, there was a significant decline in catches in Norway and Scotland, but an increase in Sweden.

In Iceland, the increased landings can be explained by
increased salmon ranching efforts. For example, in 1991, $76 \%$ of the total landings consisted of ranched fish.

Although a number of different regulatory measures to decrease harvest have been introduced in several countries, the declines in landings probably reflect reduced abundance of wild salmon stocks. In Norway, the reduction in abundance of wild salmon is greater than the catch data show because of increasing catches of escapees from fish farms in recent years. Reasons for a reduced abundance of salmon may be numerous and complex. In some countries catches are not representative of stock abundance, as in some years a large part of the salmon run takes place outwith the fishing season.

Trend analyses of grilse/salmon ratio for the two periods 1960-1990 and 1976-1990 were performed using linear regression (cf. Table 2.1.2). No significant trends could be detected.

An approach to analyse Scottish catches of 1SW and MSW salmon separately was provided. Catch statistics from Scottish salmon fisheries have been collected since 1952. Figure 3.2.1.2 shows the annual all-gear 1SW, MSW and 1 SW + MSW catches in numbers expressed as percentage difference from the 1952-1990 mean. Both 1SW and MSW groups show declines since the mid 1970s. In the case of the 1SW salmon, catches relative to the long-term mean were low throughout the 1950s and early 1960s. In the late 1960s, catches peaked since when they have declined, falling below the long term mean from 1976 and returning to levels similar to those recorded in the 1950s. In contrast, MSW salmon catches remained above the long-term mean until 1977. Since then, they have shown a steady decline, particularly the component caught between January and April. Changes over time in the catch of both 1 SW and 2 SW cannot be explained by changes in net or rod effort alone and are believed to reflect changes in abundance.

The Working Group recommends that further analyses of long-term catch data be carried out.

## Smolt counts

Time series of total or estimated production of wild smolts were available from 5 rivers. Using linear regressions they were analyzed for trends (Table 3.2.1.2). In the Burrishoole a negative trend was evident over the last 22 and 10 years, respectively. The number of wild smolts leaving the River Imsa increased over the past 10 years. There were no significant trends in the other rivers.

### 3.2.2 Escapement

Time series of total or estimated runs of adults were
available for 10 rivers. Using linear regression, trends were estimated for the entire time period for adult counts in the different rivers, and for the last 10 years of the time series (Table 3.2.2.1). The number of adults entering the River Burrishoole decreased significantly over the past 22 years. On the River Högvadsàn in Sweden the salmon run showed an increased trend over the last 38 years. This is due to mitigation, as this river has large problems with acidification. In the River Ponoy, where records are available over the past 26 years, the number of salmon ascending the river increased at a marginal level. Over the last 10 years no trends were detected in any of the rivers.

A preliminary analysis indicated significant correlations of total counts of salmon between several rivers in UK and Ireland, and between two Russian rivers. This may indicate that the forces determining adult returns to freshwater are similar within the same geographical areas. The Working Group recommends that the analysis be carried out with returns broken down to smolt and sea year-classes.

### 3.2.3 Survival indices

Significant correlations of estimated post smolt survival between hatchery reared and wild salmon from the River Imsa, and for survival of hatchery reared and wild fish to the River Burrishoole were detected. Long-time series of hatchery smolts could then be used as indices of survival for wild smolts in the same river. Table 3.2.3.1 provides analyses of time series of estimated post smolt survival of hatchery smolts from the Rivers Burrishoole, Bush, Imsa and Lagan, and of wild smolts from the River Imsa. Furthermore, similar analyses were carried out for survival of wild and hatchery reared fish back to the River Burrishoole. No significant trend could be detected in any group over the entire period of observation. However, similar analyses over the last 10 years revealed a significant decrease in post smolt survival of hatchery reared smolts from the River Lagan and a significant increase in survival of hatchery reared salmon back to the Burrishoole.

### 3.3 Western North Atlantic

Several short-term and a few long-term datasets for North American stocks were available to the Working Group. These datasets consisted of commercial and recreational catches, estimates of returns and adult counts at fishways, and smolt survival rates for Canadian and USA rivers (Anon., 1992a).

### 3.3.1 Measures of abundance

## Short-term

Counts of small and large salmon by Salmon Fishing

Area (SFA) (Figure 3.3.1.1) obtained at fishways and counting fences in Canada since 1974 are provided in Tables 3.3.1.1, 3.3.1.2, 3.3.1.3 and 3.3.1.4. Counts of small salmon at 12 of 13 fishways or fences on systems in insular Newfoundland in 1991 were down from the 1984-1989 mean. In 11 of 13 cases, the few large salmon counted in Newfoundland were below the 19841989 mean. In the Maritime Provinces, counts of small salmon were below the mean at all 3 fishways (in SFA $20,21,23$ ); counts of large salmon declined at 2 of 3 fishways. Counts of small salmon in Québec increased over the 1984-1989 mean at 2 of 4 fishways; counts of large salmon were below the mean at all 4 fishways.

The percent change between commercial landings of small and large salmon and recreational landings in Canada for 1991 is compared to recent 5 -year averages in Figure 3.3.1.2. Commercial and recreational landings of small salmon in 1991 were as much as $90 \%$ below the 5 -year average. Commercial landings of large salmon in 1991 showed similar trends, while recreational landings of large salmon in Quebec declined in 6 of 10 SFAs. The percent change in egg deposition between 1991 and previous years in rivers of many of the management areas suggest that low catches infer low spawning escapement.

Total catches in 1991 in Maine (USA) rivers with salmon runs that are primarily of wild origin were the lowest recorded in the available time series of data and $81 \%$ below the 1967-1986 average (Anon., 1992a). Similarly, the catch of 2 SW salmon of wild origin has steadily declined since 1980 (Figure 3.3.1.3). These data suggest that low catches were due to low salmon abundance in recent years.

Stock status in Canada may be summarized as follows for grouped management areas:

Labrador, East and South Coast Newfoundland (SFAs 111): The total abundance, as inferred from catch statistics and monitoring facilities, of small and large salmon in SFAs 1 to 11 was well below average.

There are several factors that may have contributed to the overall low abundance of salmon in SFAs 1-11. In northern areas, there appeared to be low egg depositions in 1984-85. In SFAs 4-11, the exceptionally low water levels and high water temperatures in 1987 may have resulted in high mortalities of juvenile salmon. Also the smolt-to-small salmon survival in 1989, 1990 and 1991 was below the survival rates observed in 1987 and 1988 for two rivers (one in SFA 9 and the other in SFA 11).

West Coast Newfoundland, Gulf New Brunswick and Gulf Nova Scotia (SFAs 12-18): Recreational catch of small salmon and returns to the counting facility at Torrent River indicated that abundance in SFAs 12-14
was about one-half of the previous 5 years. Returns of small salmon in 1991 relative to the previous 5 years were low in SFA's 15-16. In SFA 18 there were greater returns of large salmon particularly in the fall compared to the previous five year mean although less than in 1990. In SFA 16, estimates of returns of large salmon were the largest of the past 5 years.

Atlantic Nova Scotia, Bay of Fundy Nova Scotia, and New Brunswick (SFAs 19-23): Counting facility and river spawner counts indicated, with one exception, that returns of wild small and wild large salmon were lower than either those of 1990 or the 1986-1990 mean.

Gaspé, PQ (Q1-Q4): Counting facility, river spawner counts and catch statistics show that 1991 small salmon returns were below 1990 and about the same as the 198690 mean; large salmon returns were about the same as 1990 and the 1986-1990 mean.

North Shore, PQ (Q5-Q9): Counting facility and catch statistics show that 1991 small salmon returns were above average in the western part of this area and below average in the eastern part; large salmon returns were below 1990 returns but about equal to the 1986-90 mean.

Anticosti and Ungava PQ (O10-Q11): Catch statistics and spawner counts for area Q10 and catch statistics for area Q11 reveal that both small and large salmon returns were lower than 1990 and the 1986-90 mean.

## Long-term

Long-term commercial landings in Canada by province and in Greenland is shown in Figure 3.3.1.4. Abundance of salmon is inferred to have been low at the turn of the century and in the 1950s. Reduced harvests in the last 20 years are in part a result of harvest restrictions designed to increase spawning escapement to many Canadian rivers. Trends in counts of small salmon at fishways and fences and an estimate or run-size in Canada for the period 1974-1991 are shown in Figure 3.3.1.5. These data suggest that while the abundance of small salmon was generally increasing during the period 1974-1985, that trend has been reversed during the past 5-6 year period. Trends in large salmon abundance (Figure 3.3.1.6), the important contributor to egg deposition in most mainland rivers, generally show a downward trend.

In Section 5.3.1, abundance of Canadian salmon that contribute to the West Greenland fishery is estimated for the years 1983-90 ( N 1 in Table 5.3.1.2). Abundance is estimated to have ranged from about 217,000-588,0000 1SW fish destined to be 2SW during this period, peaking in 1986 and declining in recent years (Figure 5.3.1.6).

### 3.3.2 Escapement

Assessments are available for 16 Atlantic salmon stocks in Canada. Many of these stocks are newly assessed this year, as spawning targets in terms of egg requirements have now been defined for many Newfoundland rivers. Spawning requirements are determined using the following standards: 2.4 eggs $/ \mathrm{sq} \mathrm{m}$ of fluvial habitat and 368 eggs/hectare of lacustrine habitat. Annual estimates of run size and spawning escapements relative to a target spawning requirement are provided for the Restigouche (SFA 15), Miramichi (SFA 16) and Saint John (SFA 23) in New Brunswick, Margaree (SFA 18) and LaHave (SFA 21) in Nova Scotia and Conne River (SFA 11) in Newfoundland (Table 3.3.1.5) and rivière de la Trinité (Q7), Québec as in previous years (Anon 1990a). New stocks assessed include Grand (SFA 19) and Liscomb (SFA 20) rivers in Nova Scotia and, in Newfoundland, Gander (SFA 4), Terra Nova (SFA 5), Middle Brook (SFA 5) Rocky (SFA 9), Biscay Bay (SFA 9), Northeast (SFA 10) and Humber (SFA 13) rivers (Table 3.3.1.5).

Estimates of egg depositions in 1991 may have approximated (rivière de la Trinité) or exceeded (Miramichi, Margaree and Northeast) target egg requirements in four rivers. On the remaining 12 assessed rivers, egg depositions were below target requirements, some by as much as $70 \%$ below (Table 3.3.1.5). In the Miramichi River, $40 \%$ of the large salmon were repeat spawners, the highest proportion or number on record. Reasons for shortfalls in target egg deposition vary by river systems and include effects of low water levels on juvenile survival in 1987, natural cycles, low pH , and increased marine mortality.

The percent change in total egg depositions for monitored rivers in Canada during 1991 is compared to the 1986-1990 average in Figure 3.3.1.2. Egg depositions were as much as $70 \%$ below average in 12 rivers, while $10-15 \%$ increases were noted in 2 rivers.

Spawning escapements to most rivers of the western North Atlantic were generally low as inferred from low catches and counts of salmon at monitoring facilities, and estimated egg depositions (Figures 3.3.1.2 and 3.3.1.5).

In Section 5.3.1, abundance of Canadian salmon that contribute to the West Greenland fishery is estimated for the period 1983-90 (Table 5.3.1.2). The difference between estimates of total 2 SW returns in rivers (R2) and the catch of 2 SW salmon in rivers provides an estimate of the spawning escapement of 2SW salmon:

Year $\quad$ 2SW Spawners ( 000 's)

| 1984 | 92.2 |
| ---: | ---: |
| 1985 | 104.6 |
| 1986 | 131.5 |
| 1987 | 108.3 |
| 1988 | 125.9 |
| 1989 | 113.4 |
| 1990 | 119.0 |
| 1991 | 99.0 |

The target number of $2 S W$ spawners for Canada is estimated to be between 150,000-200,000 2SW fish.

### 3.3.3 Survival indices

Estimates of survival of wild smolts to 1SW returns for 5 rivers and hatchery smolts to 1 SW returns for 3 rivers in Canada are shown in Figure 3.3.3.1. Survival of hatchery smolts released in the Penobscot River (USA) to 1 SW and MSW returns to the river is also shown.

While large annual variations in survival between years is common, many stocks have exhibited trends of reduced marine survival in recent years. Survival of wild smolts from the rivers de la Trinité and Bec-Scie in Quebéc, appears to be increasing; however, wild smolt survival in Western Arm Brook, Conne River, and the Northeast River, Placentia, Newfoundland show a decreasing trend. All three of the Canadian hatchery stocks assessed exhibit flat or declining smolt survival trends. For the Penobscot River in the USA, survival of hatchery smolts to 1 SW return has remained relatively constant, while survival to return as MSW salmon has been declining for the past decade (Figure 3.3.3.1). The increasing incidence of 1 SW salmon returns to the Penobscot River in recent years may thus be explained by a decrease in survival of MSW salmon as opposed to an increase in survival of 1 SW salmon.

While poor smolt survival years are not uniformly exhibited by all stocks, it is evident that smolt survival for many stocks is lower than in previous years.

### 3.4 Summary Stock Status

### 3.4.1 North-East Atlantic Commission

The decline of catches in several countries in the NorthEast Atlantic Commission Area suggest reduced abundance of wild salmon in recent years. The Working Group examined a number of fishery-independent measures of abundance, but was unable to detect a similar pattern of decline in stocks as noted in the catch data. There were no consistent trends in smolt production or adult counts at the facilities examined over recent decades, except in the River Burrishoole where a downward trend could be detected. The fishery-indepen-
dent data series examined were few and may not be representative of national stocks.

### 3.4.2 West Greenland Commission

Although not measured precisely, it is believed that the most abundant European stocks in West Greenland originate from the UK and Ireland. It appears that the abundance of some of these stocks has declined in recent years. Similar declines in abundance have been noted in many North American stocks that contribute to the West Greenland fishery. The decline in catch and fisheryindependent measures of abundance in North America, and the decline in catch beyond the expectation that would have resulted from effort reduction in Europe, suggest there is no reason to expect that the status of stocks that contribute to the West Greenland fishery will improve in the near future.

### 3.4.3 North American Commission

The abundances of both small and large salmon, as indicated by adult returns and commercial and recreational catches, generally show a downward trend during the last 5-6 years. Similarly, spawning escapements to many rivers of the western North Atlantic were generally low, as inferred from commercial and recreational catches, adult counts at monitoring facilities, and estimated spawning escapements and egg depositions. While large annual variation in smolt survival between years is common, many stocks in the western North Atlantic have exhibited reduced marine survival in recent years. The reasons for low abundance, reduced spawning escapements, and lower smolt survival differ among river systems and include: adverse environmental conditions in homewaters (e.g., low river discharges and high water temperatures, etc. which affect smolt production), inadequate egg depositions, and increased marine mortality.

## 4 FISHERIES IN THE NORTH-EAST ATLANTIC COMMISSION AREA

### 4.1 Description of the Fisheries at Faroes

### 4.1.1 Gear and effort

The gear in use in the Faroese fishery did not change in 1991.

In recent years, the effort in the salmon fishery has continued to decline, and in the 1990/1991 season only 8 out of 13 licenses were used. The maximum permitted number of licenses is 26 . The licensed vessels were allowed to fish from 1 November to 20 December and 3 January to 12 April (one vessel below 20 GRT was allowed to fish from 1 November to 20 December and 12

January to 30 April).
Only 2 vessels started fishing early in November, but after some good catches 40 nm north-west of the Faroes during November, the remainder joined in, resulting in high effort until the Christmas closure. Due to poor weather in January, only 2 vessels went out fishing, and catches were low. Fishing effort then increased during February and early March but dropped off again in late March and April.

A total of 369 sets was fished in the 1990/91 season, $23 \%$ of the permitted maximum of 1600 .

In the 1990/91 season, no fishery took place outside the Faroese EEZ (Figure 4.1.1.1). The fishery followed the normal pattern, beginning close to the islands and moving in a north-easterly direction out to the fishery limit during the season.

In 1991, the Faroese salmon quota was bought out by various interested parties. As a result, the Faroese boat owners will receive financial compensation for the agreed salmon quota during the next three seasons (1991/92 to 1993/94). No catches will be made in the Faroes EEZ other than by the research vessel operating under the direction of the Faroese Fisheries Laboratory.

### 4.1.2 Catches and discards

The total nominal catch in the Faroes fishery in the 1990/91 season was 202 t . This is 162 t less than in the previous season and considerably lower than the catches reported for the 1981/82 to 1986/87 seasons, but similar to the catch reported in 1987/88 (Table 4.1.2.1). As in other recent seasons, good catches were taken in December, but landings were poor in January due partly to bad weather (Table 4.1.2.2). However, unlike in previous years, the landings remained poor for the remainder of the season. The catches in number by statistical rectangle for the whole season are shown in Figure 4.1.2.1.

The catch for the calendar year 1991 was only 95 t (Table 4.1.2.1). This included 13 t caught in December 1991 by the research vessel operating in the Faroes area during the 1991/92 season.

No data are available on the numbers of farmed fish taken in the fishery because appropriate data (e.g., fin
measurements or sufficient scale samples) were not collected in the market sampling programme.

Three samples of discards were collected during the fishing season and discard rates ranged from 9.9 to $16.1 \%$; the overall estimate was $14.8 \%$. This is at the high end of the range observed in the seasons 1982/83 to 1989/90 (Table 4.1.2.3). No clear trend is apparent for the time series. The discard samples in the 1990/91 season represent approximately $8 \%$ of the total landings for the fishery. However, as the samples were all taken late in the season, they may not be truly representative of the total catch. Discard sampling by month in the last two seasons indicates a decreasing trend in the proportion of discards as the season progresses. The discard rate of $14.8 \%$ should therefore be considered as a minimum estimate for the 1990/91 season.

### 4.1.3 Catch per unit effort

The catch in number per 1000 hooks (CPUE) by statistical rectangle for the whole season is shown in Figure 4.1.3.1. The CPUE values for November and December were the among the highest recorded at this time of year since 1981/82. However, the CPUE fell markedly in February and remained fairly low for the rest of the season (Table 4.1.3.1). In the 1990/91 season, no fishing took place outside the Faroes EEZ (Table 4.1.3.2).

The CPUE for the season increased in the 1988/89 and 1989/90 seasons, but fell slightly in the 1990/91 season, although these three values are the highest in the time series (since 1981/82) (Table 4.1.3.1). There is no clear relationship between the catch and the CPUE for the past 10 seasons.

The highest catch rates were observed close to the islands in November and December, and as the season progressed, the fishery moved north-east towards the fishery limit.

### 4.1.4 Biological composition of the catch

The sea age distribution of landings at Faroes by month in 1990/91 is shown in Table 4.1.4.1. These distributions have been determined from the length distributions of 1744 salmon measured during the season using the length splits shown below; the length splits for previous years are shown for comparison.

| Sea age class | Season |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
|  | $1986 / 1987$ | $1987 / 1988$ | $1988 / 1989$ | $1989 / 1990$ | $1990 / 1991$ |
| 1SW - 2SW | - | $58-59 \mathrm{~cm}$ | $57-58 \mathrm{~cm}$ | $55-56 \mathrm{~cm}$ | $57-58 \mathrm{~cm}$ |
| 2SW - 3SW | $85-86 \mathrm{~cm}$ | $83-84 \mathrm{~cm}$ | $84-85 \mathrm{~cm}$ | $83-84 \mathrm{~cm}$ | $84-85 \mathrm{~cm}$ |
| 3SW - | $102-3 \mathrm{~cm}$ | $113-4 \mathrm{~cm}$ | - | - | - |

The sea age distribution by fishing season in the Faroes is shown in Table 4.1.4.2. In the 1990/91 season, practically all the catch comprised 2SW fish (91\%), with only $1 \%$ of 1 SW fish and $8 \%$ of $3 S W$ fish. These values lie within the ranges observed in previous seasons.

The proportion of the catch in the two smallest size classes (up to 3 kg ) was lower in 1990/91 than in the previous four seasons, while the proportion in the larger categories increased. It is not known whether this reflects a change in the origin of the catch in the Faroes fishery or a change in growth rates in the sea.

No smolt age composition of the Faroes catch was obtained in the 1990/91 season.

### 4.1.5 Origin of the catch

Data on microtag recoveries in the Faroes fishery were updated for the 1990/91 season (Table 4.1.5.1). Table 4.1.5.2 shows the derivation of raising factors for the 1981/82 to 1990/91 seasons, and Table 4.1.5.3 gives the estimated total numbers of tagged fish killed in the fishery (including discards) by sea age class and the catch rates per 1000 fish tagged. Data for UK(England and Wales) are based mainly on parr tagging. The figures for previous seasons have been modified to take account of an error in the way they were estimated. For all countries the recovery rates per 1000 fish tagged were within the ranges previously observed.

A total of 135 external tags was recovered in the Faroes fishery in 1990/91 of which 116 were from Norway, 16 were from Sweden and 3 were from Scotland. Both the microtag and external tag recovery data confirm previous observations of the relative rates at which stocks from different countries were represented in the fishery.

### 4.1.6 Exploitation rates in the Faroese fishery

The estimates of extant exploitation rates in the Faroes fishery on several stocks from Norway, Sweden and UK(Scotland) are summarised in Table 4.1.6.1.

The exploitation of hatchery stocks from the Drammen (Norway) and Lagan (Sweden) have shown similar changes with levels being quite low in the 1986/87 and 1987/88 seasons but higher in 1985/86 and in the two most recent seasons. The two Norwegian hatchery stocks (Drammen and Imsa) showed opposite trends, with the exploitation rate on the Drammen stock falling in 1990/91 after a 2 year peak while that on the Imsa stock rose after a 3 year trough. The exploitation rates on wild fish from the Imsa and North Esk have been very much lower in the past 5 years than previously although there was a slight rise for 2SW fish in 1990/91. There is no clear relationship between the trends for individual stocks and the catches recorded in the fishery.

### 4.2 Description of Homewater Fisheries

### 4.2.1 Gear and effort

No changes in the regulations affecting salmon fishing gear in 1991 were reported for any countries except Norway and Scotland. In Norway, the use of monofilament nets was banned for catching anadromous salmonids. In Scotland there were changes in the regulations affecting gear or fishing period for rod and line fisheries in four rivers.

Fishing effort was thought to have been reduced in France, Ireland, UK (Northern Ireland), UK (England and Wales), UK (Scotland), Russia and Sweden. Factors affecting this reduction are thought to have included perceived reductions in stock abundance and weather conditions (e.g., early freezing in Russia and low river flows in most other countries). The following additional changes in effort were also reported:

France: On the lower parts of the main rivers of Brittany and Normandy the angling season was extended for one month for fly fishing only.

Iceland: The set net fisheries on the Hvita in Borgarfjordur were rented and thus not operated.

Russia: The operation of several net fisheries was modified to increase the escapement.

UK (England and Wales): Short-term netting restrictions were in operation on four rivers. The closed area around the River Esk (Yorkshire) was enlarged, reducing the area available to beach nets.

### 4.2.2 Catches and catch per unit effort

Revised estimates of total nominal catches of salmon by country were available for the 1990 fishery and provisional estimates for the 1991 fishery (Table 2.1.1). Catches of ranched fish and fish farm escapees are included in these statistics. Data for 1991 for Ireland, Norway and UK (Scotland) are incomplete.

The provisional total catch for the 1990 fishery was only $56 \%$ of the final declared catch. Accordingly, the provisional 1991 catch of 2826 t for the North-East Atlantic Commission area is likely to increase considerably. The 1991 figure is much lower than the averages for the previous 5 and 10 years ( 4,735 and $4,911 \mathrm{t}$, respectively), reflecting decreases in most countries, especially France, Ireland, UK (Northern Ireland), UK (England and Wales), UK (Scotland) and Russia, although catches for Finland, Iceland and Sweden were above the 10 -year average. Specific information was provided as follows:

Finland: the 1991 catch was the highest since the mid1970s; this is believed to be partly attributable to the coastal netting restrictions in place in Norway.

France: The catch was similar to the 1990 catch but was believed to be limited by later entry of fish to freshwater and poor fishing conditions due to drought.

Iceland: The catch was over twice the 10 year average, reflecting the increasing contribution of ranched salmon (estimated at $76 \%$ ) to the fisheries. Recent improvements in marine survival of ranched fish have also contributed to the increased catch.

Norway: Catches in 1991 were low, as in the previous two years, probably reflecting reduced effort due to the management measures in coastal fisheries.

Russia: Catches were lower than any in the previous 30year period. This largely reflects the reduced exploitation rates on several stocks.

UK (England and Wales): The low catches (57\% of the 5-year average) were influenced by prolonged low flows and a reduction in effort, partly as a result of regulatory changes and partly in response to poor catches.

UK (Northern Ireland): The provisional catch of 55 t was the lowest on record and was only $53 \%$ of the previous 5 year average. Reduced catches generally reflected the poor conditions for drift netting and reduced fishing effort resulting from perceived low of abundance of fish.

UK (Scotland): The total nominal catch in 1990 was $30 \%$ less than that in 1989 and the provisional 1991 catch showed a further considerable decline. Fishing effort has been reduced as a result of the buy-out of netting stations and also in response to perceived low stock abundance.

Catch-per-unit-effort data were available for various salmon fisheries in Wales [UK (England and Wales)]. These indicated that in 1991, CPUE in most net fisheries was reduced relative to average values for the period 1986-90. Catch-per-unit-effort data were also available for rod fisheries in the River Conwy. Average CPUE in 1991 ( 0.015 salmon per hour) was lower than the range for the previous 5 years $(0.020-0.045)$ but not as low as in 1984 (0.012), which, like 1991, had been a very dry year. No other CPUE data were available.

### 4.2.3 Composition of the catch

Reported national salmon catches by sea age for several countries in the North-East Atlantic Commission area are summarised in Table 2.1.2. These data include estimates derived from scale sampling and from the weight
composition of the catch. Specific information on age composition was provided as follows:

France: The proportion of 1SW fish (39\%) fell within the range for the last four years ( $29 \%-77 \%$ ) but was lower than the average ( $50 \%$ ).

Iceland: In 1991, grilse in the fishery were smaller than usual and males predominated.

Norway: The proportion of 1SW fish in the catch by numbers ( $65 \%$ ) was similar to 1990.

Russia: As in 1989-90, 1SW fish were the predominant age group ( $70 \%$ ). The increase in this proportion from earlier years is a result of a ban on the Pechora river fishery, which formerly contributed $50 \%$ to the total Russian catch and where MSW fish predominated.

UK (England and Wales): The proportion of 1SW fish in the catch in 1991 ( $61 \%$ ) was lower than in previous years, suggesting poor marine survival of the 1990 smolt year class. No reliable data are yet available for 1991.

UK (Scotland): The proportion of 1SW fish in the catch fell from an average of $60 \%$ in 1986-89 to $48 \%$ in 1989 reflecting the apparent poor survival of the 1989 smolt year class. The slight increase to $54 \%$ in 1991 was caused by the reduction in MSW catches not a recovery in 1SW catches.

### 4.2.4 Origin of the catch

Table 4.2.4.1 indicates the origin of the salmon catches in each country based upon recoveries of tags over a number of years. Double crosses indicate the principal component of the catch and single crosses other significant contributions. Rare recoveries of one country's tags in another country are indicated by dashes and were assumed to indicate very minor contributions to catches. It is apparent that there is normally a pattern of interchange between neighbouring countries, although this exchange may not always be even. It must be noted that this table reflects the relative size of national stocks. Countries with small stocks are unlikely to contribute significantly to a fishery targeting mainly other larger stocks. The table does not indicate the proportion of fish originating from a given country that are caught in other countries.

Table 4.2.4.2 shows estimated contributions of ranched and farmed fish to national catches. In this context, ranching is defined as the release into the wild of reared smolts with the intention of attempting to harvest all returning adults. Releases of reared fish to enhance wild stocks or compensate for lost wild production are therefore ignored.

The only country in the North-East Atlantic Commission area known to be ranching in this way is Iceland, where ranched fish comprised $76 \%$ of the catch in 1991. However, in France there is a small experimental ranching exercise. In addition, 14 t of the catch in Sweden comprised fish that have been released for mitigation purposes, but are not expected to contribute to wild spawning populations.

The only countries in which farmed fish are thought to make a significant contribution to fisheries are Norway and UK (Scotland). In Norway, where extensive surveys have been undertaken since 1988, farmed fish appear in both marine and freshwater fisheries. Estimates of the proportion of farmed fish in various Norwegian fisheries were highly variable between sites but indicate that the proportion of farmed salmon was much lower in samples taken in fresh water than in coastal areas. The proportion of farmed fish in the catch seems to have been relatively constant in the period 1989-91.

In UK (Scotland), sampling in 1990 indicated that most of the reared fish caught in fisheries had escaped or been lost from sea cages. In 1991, however, sampling on the west coast revealed that most of the farm-origin fish were derived from losses or releases of smolts or parr. On the east coast, where the incidence of farm escapees was low, most of the farm origin fish were adult escapees.

In all other countries, farmed fish are thought to form only a very minor (or negligible) part of the catch.

### 4.2.5 Exploitation rates

Estimated exploitation rates in homewaters for monitored stocks in Ireland, UK (England and Wales), UK (Northern Ireland) and UK (Scotland) are summarised in Table 4.2.5.1 and similar data from Iceland, Norway, Russia and Sweden are shown in Table 4.2.5.2. The patterns observed for different stocks are very variable. Exploitation on the River Drammen and Lagan stocks (hatchery reared fish) were higher than average in 1991 while the rates for the North Esk (UK (Scotland)) and Imsa (Norway), and for hatchery-reared fish on the River Bush (UK (Northern Ireland)), were lower. For most other stocks (including wild fish from the River Bush) rates were similar to those estimated for 1990. On the Russian rivers, fishing traps are operated every day and the exploitation rates are adjusted by altering the proportion of days on which the catch is released or killed. Exploitation rates were reduced in 1991 to protect spawning stocks.

### 4.2.6 Effects of recent management measures in Norway

Full details of the management measures introduced in

Norway in 1989 are given in Anon. (1990a), Appendices 2 and 3.

The impact of the measures on catches in Norwegian homewaters in 1989 and 1991 is shown in Table 4.2.6.1. In the period 1982-1988, the total nominal catch of salmon fluctuated between 1,076 and $1,623 \mathrm{t}$. It decreased to 905 t in 1989, 930 t in 1990 and 885 t in 1991, probably as a result of the new management measures. In 1989, 1990, and 1991, the marine catches of salmon were $488 \mathrm{t}, 514 \mathrm{t}$ and 471 t , respectively, which is much lower than for 1982-1988, when this catch varied between 841 t and $1,324 \mathrm{t}$. The catch in the marine salmon fisheries, excluding drift netting, was close to the average for this period.

It is likely that the ban on drift netting in 1989 has resulted in a larger number of salmon being available to the other marine homewater fisheries. The additional regulation of these fisheries has probably resulted in a substantial increase in freshwater escapement suggested by increased catches in freshwater. In 1989, 1990 and 1991, the freshwater catch accounted for $46 \%, 45 \%$ and $47 \%$ of the total nominal catch respectively, compared to between 18 and $27 \%$ over the years 1982-1988. Increased freshwater escapement is also suggested by the reduction in marine exploitation rates on most components of the River Imsa salmon stock during 1989-1991. This was not the case for salmon of the River Drammen stock, however, because drift net exploitation on this stock has always been low.

The salmon fishery on the west coast of Norway intercepts stocks from Finland, Russia and the Swedish west coast on their return to their home rivers. Exploitation in Norway on 1SW fish tagged as smolts on the River Lagan in 1989, 1990, and 1991 was lower (average 1\%) than in 1985-88 (average 7\%) (Table 4.2.6.2). This suggests that the management measures introduced in Norway in 1989 also benefitted Swedish west coast stocks.

The frequency of net-marked salmon entering a river may also give information about changes in netting effort on the migration route. The proportion of net-marked salmon has been recorded in several Norwegian rivers since 1978. In most of these rivers, sampling took place from 1978 to 1986 and was then re-established in 1990 and 1991. Table 4.2.6.3 shows un-weighted means of the proportion of net marked salmon in angling catches from 12 rivers in the period before the extensive homewater regulations were introduced, and the un-weighted means of the proportion of net-marked salmon in the same rivers in 1990 and 1991. In all except one river, the proportion of net-marked salmon recorded in 1990 and 1991 was much lower than the un-weighted means during the period 1978-1988. The reduced proportion of net-marked fish may be accounted for by the manage-
ment measures introduced in the Norwegian homewater fishery in 1989.

### 4.3 By-Catches of Fish, Birds and Mammals in Drift-Net Fisheries

Drift-net fisheries are currently operated by six countries in the NEAC area; France, Finland, Ireland, Norway, UK (England and Wales) and UK (Northern Ireland). These fisheries also target migratory trout (Salmo trutta), which are not, therefore, regarded as a by-catch.

France: Drift-nets are used in two estuarine fisheries targeting a variety of fish species. There is no information to indicate that any by-catches occur in these fisheries.

Finland Drift-nets are operated in freshwater in the Tana River. Seals (Phoca vittulina) are reported to have been caught in this fishery very rarely. Pike have also been recorded as by-catch. It should be noted that there is no information for the Norwegian portion of this fishery.

Ireland: Drift-nets are used extensively around the coast of Ireland and account for $70-80 \%$ of the reported salmon catch. There is very little information available on the by-catch of other species but this is considered to be extremely small. Species which may become entangled include bass (Dicentrarchus labrax), guillemots (Uria algae) and harbour porpoises (Phocoena phocoena). Recent reports indicate that the level of seal predation on salmon in drift-nets has increased in recent years, but despite this, there have been no reports of seals becoming entangled in the nets. The incidence of porpoise capture is extremely low although these animals are regularly recorded in Irish waters.

UK (England and Wales): A total of 328 salmon driftnets is used to fish for migratory salmonids in England and Wales; all nets are less than 550 m in length and about half of them target almost entirely migratory trout. Data on by-catches are only available for the fishery on the north-east coast of England, where 143 nets are used.

Very few fish other than salmonids are caught by these nets, but small numbers of mackerel are taken, and other species including bass and grey mullet may be seen very occasionally.

Despite frequently being seen around nets, grey seals (Halichoerus grypus) are rarely caught in the drift nets because they can break the netting. Harbour porpoises occasionally become entangled in the drift-nets, but because the fishermen have to remain in attendance of the gear they are normally able to release them before they are seriously harmed. Scientific observers worked
on four vessels in the Northumbria Region throughout the 1976 and 1977 season, and on three in the Yorkshire Region throughout the 1979 season. These observers saw four porpoises entangled in nets (three in 1976 and one in 1979) all of which were released alive.

Drift-nets rely upon being relatively difficult to see in order to catch fish, and for this reason diving birds can become caught. Guillemots, razorbills (Alca torda) and very occasionally puffins (Fratercula arctica) are caught. Once again, the netsmen endeavour to remove them alive. No data are available on the numbers killed.

By catches as a result of "ghost fishing" are not a problem in the English drift-net fishery, where nets are rarely lost or abandoned.

UK (N.Ireland): A total of 117 drift-nets were licensed in Northern Ireland in 1991. No quantitative information on the by-catch is available for this or any other year, but it is known that porpoises and guillemots are occasionally caught.

## 5 FISHERIES RELEVANT TO THE WEST GREENLAND COMMISSION

### 5.1 Description of the Fishery at West Greenland, 1991

In 1991, the fishery at West Greenland (NAFO subarea 1) was opened on 5 August and ended in November, although the official closing date was 31 December. The total nominal catch was 437 t (Tables 5.1.1 and 5.1.2), which is 210 t more than in 1990, when the total landings were 227 t .

The TAC for 1991 was set unilaterally at 840 t , and divided into a "free" quota of 373 t and a "small boat" quota of 467 t . Because the landings in 1991 were small, this arrangement had no practical implications.

The geographical distribution of the fishery in 1991 (Table 5.1.2) differs little from previous years. The landings in NAFO Divisions 1A and 1B were again low, whereas, landings in NAFO Divisions $1 \mathrm{C}, 1 \mathrm{E}$ and 1 F were high, with highest figures in Division 1F.

The salmon fishery in Greenland is a small boat fishery and is executed in inshore and coastal areas. Approximately $80 \%$ of the total landings were taken by boats smaller than 30 feet. No information on effort is available for 1991, but the landings during the two first two weeks is given for 1980 to 1991 (see text table below).

The nominal landings during the first two weeks, 1980-1991 (in tonnes).

| Year | First week | First two weeks |
| :---: | :---: | :---: |
| 1980 | 260 | 711 (1-14 Aug) |
| 1981 | 465 | 735 (15-28 Aug) |
| 1982 | 470 | 766 (25 Aug - 7 Sep) |
| 1983 | 105 | 192 (10-23 Aug) |
| 1984 | 17 | 58 (10-23 Aug) |
| 1985 | 204 | 361 (1-13 Aug) |
| 1986 | 509 | 848 (15-28 Aug) |
| 1987 | 439 | 737 (25 Aug - 7 Sep) |
| 1988 | 219 | 337 (25 Aug - 7 Sep) |
| 1989 | 131 | 219 (18-31 Aug) |
| 1990 | 12 | 38 (1-14 Aug) |
| 1991 | 114 | 191 (5-18 Aug) |

### 5.1.1 Composition and origin of the catch, 1991

The Working Group examined the composition and origin of salmon caught at West Greenland based on discriminant analysis of scale samples from NAFO Divisions 1C, 1D, 1E. The database used to develop the discriminate function consisted of 788 North American and 788 European known-origin salmon collected at West Greenland from 1980 to 1991. There were insufficient numbers (36) of salmon of known-origin identified in the 1991 sampling to develop a discriminant function solely on the 1991 samples. Three discriminant functions were used - one for salmon with river age 1 , one for salmon with river age 2 , and one for salmon with river age 3 and older.

Scale samples which were independent of the discriminant analysis database and weighted to 1991
river-age distributions at West Greenland were used to test the discriminant functions. The results indicated an overall misclassification rate of $19.5 \%$ and error rates of $\pm 3.3 \%$, based on prior probabilities of 0.5 . This database and the discriminant function were accepted by the Working Group for examination of the 1991 West Greenland fishery. The Working Group expressed concern over the lack of suitable test samples of known origin salmon collected from the Greenland fishery in 1991. Scale samples should be collected from 2SW salmon in home waters in 1992 and used as a database to recalculate the continental proportions in 1991.

The results of classifying salmon in samples from commercial catches in 1991 indicated that the North American proportion was $65 \%(95 \% C L=69,61)$, and the European proportion was $35 \%$ ( $95 \% \mathrm{CL}=39,31$ ) (Table 5.1.1.1). In 1991, $52 \%$ of the catch was taken at the same time as the samples compared to $26 \%$ in 1990 , $73 \%$ in $1989,62 \%$ in 1988, and $85 \%$ in 1987. Although the per cent coverage improved over 1990, the catch samples are not as representative as those taken during to 1987-89. In 1991, $3.1 \%$ of the catch was sampled for determination of continental proportions compared to $5.8 \%$ in $1990,8.0 \%$ in $1989,3.8 \%$ in 1988 and $4.5 \%$ in 1987. This suggests that the samples may be representative of catches taken during the same period, but may not be representative of the entire fishery. Table 5.1.1.1 shows the variability in the proportions of North American and European salmon in the fishery since 1969.

An alternative estimate of the overall proportion of North American and European-origin salmon for the years 1982-1990 was derived by weighing NAFO division samples by catch in numbers. Information from the nearest NAFO Division was applied to divisions with no samples. The table below gives the results:

| Year | Weighted by catch in numbers |  |  |  | Percentage of all samples combined |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | NA |  | EU |  | NA | EU |
|  | \% | Wt (t) | \% | $\mathrm{Wt}(\mathrm{t})$ |  |  |
| 1982 | 57 | - | 43 | - | 62 | 38 |
| 1983 | 40 | - | 60 | - | 40 | 60 |
| 1984 | 54 | - | 46 | - | 50 | 50 |
| 1985 | 47 | - | 53 | - | 50 | 50 |
| 1986 | 59 | 537 | 41 | 423 | 57 | 43 |
| 1987 | 59 | 556 | 41 | 411 | 59 | 41 |
| 1988 | 42 | 439 | 58 | 544 | 43 | 57 |
| 1989 | 55 | 179 | 45 | 158 | 56 | 44 |
| 1990 | 74 | 168 | 26 | 59 | 75 | 25 |
| 1991 | 63 | 267 | 37 | 170 | 65 | 35 |

In 1991, the estimated number of fish caught was 102,966 from North America and 60,934 from Europe for a total of 163,901 . The time series of catches by number, 1982-1991, is provided in Table 5.1.1.2.

As in previous years, there were no temporal or spatial trends in the proportions of fish by continent; significantly higher proportions of North American salmon were observed in Divisions 1C and 1D than in 1E:

| NAFO <br> Div. | Nominal <br> catch (t) | \% N. American <br> origin | \% European <br> origin |
| :--- | :---: | :---: | :---: |
| 1A | 22 | $*$ | $*$ |
| 1B | 21 | $*$ | $*$ |
| 1C | 109 | 67 | 33 |
| 1D | 31 | 71 | 29 |
| 1E | 106 | 52 | 48 |
| 1F | 148 | $*$ | $*$ |
|  |  |  |  |
| * not sampled |  |  |  |

Information on country of origin can be derived from recoveries of tags (both Carlin and coded-wire) at Greenland in 1991. Salmon landings at Greenland were again scanned in 1991 for adipose finclips and codedwire tags (CWTs) using procedures similar to those in previous years.

In 1991, a total of 7,357 salmon ( $4.5 \%$ of the West Greenland catch) was examined for adipose finclips and CWTs by Canadian, USA, and Danish scientists. In the sample, 121 ( $1.64 \%$ ) had adipose finclips, and CWTs were recovered from 37 ( $30.6 \%$ ) of the finclipped fish (Table 5.1.1.3). Thus, the overall proportion of the catch sample that had CWTs was $0.50 \%$, compared to $0.84 \%$ in 1990, $0.64 \%$ in 1989, $0.50 \%$ in 1988, and $0.58 \%$ in 1987. The proportions of fish having adipose fin clips and CWTs sampled at each port did not differ from a uniform distribution ( $\chi^{2}=2.8$ and 0.4 , respectively). Thus, in 1991, the distribution of adipose fin clips and microtags were evenly distributed throughout the fisheries, similar in occurrence to the distribution in 1988. In most other years since 1985 the proportions of the fish having adipose fin clips and CWTs decreased in the fishery from north to south. This pattern particularly holds true for North American-origin tagged salmon; whereas European salmon appears to generally be more evenly distributed throughout the NAFO Divisions (Table 5.1.1.4).

In 1991, 37 CWTs were recovered in 1991 and apportioned by country as follows: (70\%) USA, (5\%) Canada, (5\%) Ireland, (8\%) England and Wales, (5\%) Iceland (3\%) Scotland, and (3\%) from Northern Ireland (Table 5.1.1.5). All of these tags came from 1SW salmon, most having been released as hatchery-reared
smolts in 1990; although two were tagged as wild smolts in England and Wales, in 1990. Recoveries of tagged salmon from Canada and European countries are too few in number to identify differences in distribution in the fisheries. However, the proportions of USA-origin microtagged fish were evenly distributed throughout the fisheries sampled.

The contribution by various countries to the 1991 West Greenland harvest cannot be determined at this time due to differential survival of stocks tagged, as well as the proportion of coded-wire tagged fish relative to total smolt production in each country.

Carlin tags (13) of USA origin were recovered in the 1991 West Greenland fishery; further information is provided in Section 5.1.3. Carlin tags of other national origin are known to have been recovered at West Greenland, but data are not available at this time.

The Working Group considered an estimate of the number of Maine salmon harvested at West Greenland in 1991 using the proportional harvest method (Anon., 1989a). In this method, the number of 1SW North American salmon of river age 1 in the West Greenland fishery are apportioned by the relative production of age 1 smolts by Maine and Canadian hatcheries the year before. The contribution to the Greenland harvest of other US origin (i.e., Merrimack, Connecticut) smolts was considered low. The proportional harvest method was used to provide a current year estimate of the harvest.

Input data for the method are shown in Tables 5.1.1.6 and 5.1.1.7. The estimate of the North American harvest given in Table 5.1.1.7 differs slightly from that given in Table 5.1.1.4 because it is derived in a different way. The difference is small, however, and in order to maintain consistency in the method between years, the estimate has not been adjusted. Based on this method, the estimate of Maine salmon harvested at Greenland in 1991 was 3,757 (Table 5.1.1.8). From 1978 to 1991 the harvest estimates of Maine-origin salmon in the West Greenland fishery ranged from 2,350 to 33,351 salmon. With the exception of 1978 and 1982 the standard error around the estimate was less than $10 \%$ of the mean and averaged about $7 \%$ of the mean. A sensitivity analysis was conducted of the harvest estimate and indicated that most of the variability in the harvest estimate comes from the variation in the estimated number of river age 1 salmon of North American origin. Previously, the Working Group (Anon., 1986) noted that the tendency of scale readers would be to classify river age 1 salmon as river age 2. Therefore, an increase in precision of the classification would probably result in an increase in the harvest estimate.

The Working Group reviewed information on the use of
a "neural network" as an alternative technique to determine the continental origin of salmon caught at West Greenland from scale characters. The same data sets were used to develop and test the discriminant function and a neural network, and the results were compared. The neural network was set up with the four input variables (river age, age, fork length, and scale circuli counts) separated into discrete classes in order to provide 33 input neurons; the 2 output neurons corresponded to North American and European; and the network was given 17 hidden layer neurons. The neural network gave a more accurate classification ( $85.8 \%$ correct) of samples of North American and European groups than the discriminant analysis ( $80.3 \%$ correct). The Working Group concluded that the neural network technique looks promising and may offer an alternative to the discriminate analysis. Its application should be further investigated. The neural network has the advantage that it can use a larger number of input variables. Several disadvantages were noted: there is some subjectivity in the data pre-processing; learning algorithms vary between software implementations, and learning is sensitive to ordering of the data and tolerance settings.

The Working Group recommends that the neural network methodology undergo further testing using simulation data sets and run parallel with discriminant function analysis. Specifically, the Neural Network should be evaluated with respect to its:

1. sensitivity of classification success with respect to the order of input observations,
2. sensitivity of classifications success to category boundaries applied to the input parameters,
3. ability to discriminate simulated samples drawn from underlying distributions with known parameters,
4. relationship to traditional statistical methods, particularly log linear models for categorical data.

### 5.1.2 Biological characteristics of the harvest

Biological characteristics (length, weight, and age) were recorded from samples of commercial catches from NAFO Divisions 1C, 1D, and 1E in 1991 using the results of discriminant analysis to divide samples into North American and European components. A summary of these data is provided in Table 5.1.2.1.

As previously observed, North American 1SW salmon were significantly shorter and lighter than their European counterparts, both overall and on an individual NAFO Division basis. Two sea-winter salmon of North American origin were not different in length but were lighter than European-origin salmon both overall and between

NAFO Divisions at the $5 \%$ level of significance.
The sea age composition in 1991 (Tables 5.1.2.2 and 5.1 .2 .3 ) of $94.7 \% 1 \mathrm{SW}, 4.9 \% \mathrm{MSW}$, and $0.3 \%$ previous spawners indicated that there were proportionately fewer 1SW salmon and more MSW salmon than in 1990. In 1991, the 1SW components for both North American ( $95.6 \%$ ) and European ( $93.4 \%$ ) salmon were lower than their respective components in 1990.

Based on $63 \%$ North American salmon by number in the 1991 West Greenland salmon catches, the catch at age by continental origin was as follows:

| Sea age | NA | EU | Total |
| :--- | ---: | ---: | ---: |
| 1 | 98,126 | 56,972 | 155,098 |
| 2 | 4,285 | 3,859 | 8,144 |
| PS | 602 | 104 | 706 |
| Total | 102,966 | 60,935 | 163,901 |

The mean smolt age of salmon of North American origin has varied more than that of European fish (Table 5.1.2.4). There are no trends in the mean smolt ages of European-origin salmon between 1968-1991. The proportion of North American-origin river age 1 salmon has been increased steadily from $2 \%$ in the 1986 samples to $8.8 \%$ in the 1990 samples. In 1991, it decreased to $5.2 \%$. In 1991, samples ( $<1.0 \%$ ) of salmon thought to be fish farm escapees were found in the Greenland catches. The decrease in numbers of North American salmon of river age 4 years and older from the mean value of $22.0 \%$ from $1968-90$ to $17.8 \%$ in 1991 suggests that either production or migration of salmon from the northerly portion of the range in North America has decreased.

### 5.1.3 Historical data on tag returns and harvest estimates

There are three main methods for estimating the harvest of USA-origin salmon in the fishery at West Greenland (Anon., 1989a). One of these methods, the proportional harvest model (updated in Section 5.1.2) provides estimates for the current fishery year. Two other approaches, the Carlin tag harvest model and the CWT harvest model, rely on the fraction of tags in the homewater run in the following year. Therefore, the Carlin and CWT methods can only provide estimates in the year after the fishery.

## Carlin tag method

The parameters in the Carlin tag harvest model for 1SW salmon remain as reported in the previous assessment
(Anon., 1991a). The 2SW harvest estimates were not updated this year as there were no 2SW tags (Maineorigin) returned from the fishery in 1991. All 1SW returns in year i are raised to harvest estimates with the ratio of tagged to untagged 2SW returns in homewaters in year $\mathrm{i}+1$ (RATIO). RATIO values are shown in Table 5.1.3.1. For 1991, the estimates of tags and run size were 29 tags and 2057 fish, respectively. The relatively low numbers of tags in the run resulted in the 2nd lowest RATIO parameter used in the harvest model time series (RATIO for $1991=0.0138$ ). This means that each tag reported accounts for a large number of fish. Only harvest for an assumed passage efficiency of $85 \%$ is presented.

A number of tags (30) were returned in 1991 from Greenland without information as to year of recapture. These tags were presumably caught mostly as 1 SW fish in the year following release. These tags were not developed into estimates of harvest however as they are assumed to form part of the pool of "unreported" tags for previous years and are already accounted for in the harvest estimate by the reporting rate factor.

The updated time series of tag returns from Maine-origin 1SW salmon in West Greenland can be found in Table 5.1.3.2. Tag returns (to date) for the 1991 fishery total 13 tags with the largest recovery occurring in NAFO Division 1C. Estimated harvest of 1SW salmon in West Greenland is summarized by year for $85 \%$ passage efficiency in Table 5.1.3.3. The harvest estimates for the 1990 fishery totalled 1,525 salmon at $85 \%$ efficiency and are primarily distributed in NAFO Divisions 1B to 1 E . This estimate of 1525 fish in 1990 is about $60 \%$ lower than the estimate of harvest for 1989 (which was the highest recorded) and about $17 \%$ lower than the previous 10 -year mean of 1843 fish.

## CWT method

In 1990, CWTs from USA-origin salmon, including fish from the Connecticut, Merrimack and the Penobscot rivers, were recovered in West Greenland. Using the methodology in Anon. (1988a), the Working Group estimated harvests based on the CWT sampling programs at both West Greenland and in home waters. Ratios of CWT tagged to untagged 2SW salmon (RATIO) returning to the Connecticut, Merrimack and Maine rivers in 1991 were $0.757,0.210$ and 0.213 , respectively (Table 5.1.3.4).

As not all weeks or areas of the fishery were sampled, assumptions concerning the numbers of tags caught in those weeks and areas were again necessary (Anon., 1989a). In 1991, the Working Group recommended that the various stratification procedures of the CWT sampling be investigated (Anon.,1991a). Three stratification strategies were considered, each producing different
levels of harvest for the three stock groups (Maine, Connecticut, and Merrimack):

1. area stratification with no stratification over time.
2. area stratification with time stratified into early and late segments.
3. area stratification with the highest resolution time stratification that the data would allow.

Variability in the harvest estimate generally increased with the number of strata involved in the estimates and, as well, with increased number of strata, the percentage of the catch examined decreased, in some cases down to only $1 \%$. The Working Group agreed to adopt the following criteria concerning post-stratification strategies which would be used to develop harvest estimates: 1) the strategy with the maximum number of strata that does not result in an increase in the estimate variability;and 2) the strategy that pools sampling units with adjacent units so that no individual stratum has less than $2 \%$ of the catch examined. Following these criteria, a post-stratification strategy was selected for the Maine stocks and then the same strategy was applied to the estimate for the Merrimack and Connecticut stocks. Harvest estimates for the 1SW component of these three stocks for 1990 at West Greenland were 231 for Connecticut, 1070 for Merrimack and 1613 for Maine. Past estimates of harvest and the coefficient of variation are given in Table 5.1.3.5. Estimates for the Connecticut and Merrimack stocks in 1990 were the highest of the four year series while for Maine the estimates were lowest. The various estimates of harvest are shown graphically in Figures 5.1.3.1 and 5.1.3.2.

The harvest of Maine-origin salmon at West Greenland decreased in 1990 from the previous year, as estimated by the three methods (Figure 5.1.3.1). The proportional harvest method in 1989 and 1990 provided estimates of harvest of Maine-origin fish which were significantly higher than the CWT method (Figure 5.1.3.2). As escapees from North American aquaculture facilities could increase the estimates provided by the proportional harvest method, the Working Group recommended further investigation of this possible explanation of the discrepancy between the two methods.

### 5.1.4 Patterns of stock composition in the harvest

The North Atlantic Salmon Working Group has been requested with respect to Atlantic salmon in the West Greenland Commission area to examine historical data on stock composition for the presence of predictable patterns (trends) in the proportion of North American and European origin salmon in catches at West Greenland.

The continental proportions of Atlantic salmon caught at West Greenland for NAFO Divisions 1A to 1 F combined have already been shown in Table 5.1.1.1. The estimates are divided by research vessel catches (1969 to 1982) and commercial catches (1978 to 1991). The method of assessment is outlined in Section 5.1.1. A higher proportion of North American salmon has been indicated from the sampling program for recent years (1989-1991).

Distribution by continent and country as inferred by microtag return rates

The recovery rates of microtagged salmon from the West Greenland fishery per 1,000 fish examined and the continental distribution in each year is shown in Table 5.1.1.4 and Figure 5.1.4.1.

The proportion of tags recovered per 1,000 fish examined is generally higher for salmon of North American origin than for European salmon. Similar proportions were observed in 1986 for NAFO Divisions 1B, 1D and 1E in 1986, 1E in 1989 and a higher European component was noted in the recoveries in 1E in 1990. Two distinct north-south trends were observed for the N . American data in 1989 and 1990. A less distinct trend was noted in 1986. This trend was not consistent over the other years or for the distribution of European tags in any year.

The distribution of recaptures and recovery rates of microtagged salmon from the different countries for the period 1986-1991 is shown in Table 5.1.4.1. For North American origin fish the general trend appears to indicate a higher proportion of microtagged fish in the north decreasing to the south. This trend may also apply to fish of Irish origin. No clear pattern emerges for fish from England and Wales and insufficient returns have been recorded for other countries to indicate differences in distribution.

## Trend analysis of the stock composition

The data on stock composition used in the analysis came from the reports of the Working Group on North Atlantic Salmon (Anon. 1991a). It consists of the North American proportion landed by NAFO division by week, 19831990. Results for European proportions can be inferred from the proportion of the North American component of the catch.

Trend analysis of the proportional data was carried out on arcsine-transformed data using SAS Procedures for General Linear Models and Regression. Residuals were examined by plots of residuals on year, on predicted values, on North American proportions and on rank of the dependent variable. Analyses of combined catch and proportional data was done by multivariate analysis of
variance. Trend analysis was completed using the standardized coefficients from the ANOVA.

Table 5.1.4.2 shows the results of trend analysis for data collected from 1986 to 1990. Analysis of temporal trends shows that out of 15 series analyzed only 2 were significant. This result could be anticipated by chance alone. The spatial analysis shows that out of 25 series analyzed, 14 were significant; a much higher number than would be expected by chance alone. Further analysis is warranted on these trends.

ANOVA was used to analyze a model of proportional data with year, week and NAFO division as covariates. The North American proportion was significant at less than $5 \%$ only for the year effect (Table 5.1.4.3). The North American proportion was not significantly affected by week or NAFO division effect. The standardized coefficients show the relative magnitude of the annual differences in the North American proportion (Figure 5.1.4.2). Not only have there been differences between years, but, when tested for temporal trends by regression analysis the standardized coefficients showed an increasing linear trend with the highest proportion occurring in 1990. By inference it can be assumed that the proportion of European origin salmon in the catch has decreased. This trend was also noted in the analysis of the continental proportions of salmon in the sampling program.

The significant YEAR/NAFO effect (Table 5.1.4.3) indicates that the spatial patterns of catches over years are variable so therefore no predictable annual pattern of catch by NAFO division can be shown.

The recoveries of micro-tagged salmon indicated a northsouth trend for tagged North American stocks in some years, with greater numbers in the Northern NAFO Divisions. This trend was not as evident for the distribution of European tags. Analysis of proportional conti-nental-origin derived from scale characters indicated to consistent north-south distribution of North American or European components. However, in recent years there was an increase in the North American Component at West Greenland.

### 5.2 Description of Homewater fisheries

## European homewater fisheries

Tagging experiments have demonstrated that all countries listed in the National Catch Tables (2.1.1.1) contribute fish to the West Greenland fishery.

However, stocks from these countries contribute to the fishery to differing extents, both because the proportion of MSW salmon in the stocks varies and because of differences in their migratory behavior in the sea. Although the relative contributions have not been
estimated precisely, MSW stocks from UK, Ireland and France are thought to contribute to the fishery at a higher rate than Scandinavian stocks.

MSW salmon stocks have been in decline in many parts of Europe for at least the last 20 years. The extent of the change varies, but catches in some rivers which used to support mainly MSW salmon are now mainly 1SW fish [e.g., Rivers Exe and Eden in UK (England and Wales)].

The closure of the Norwegian drift net fishery has had beneficial effects on other fisheries in Norway, Finland, Russia, and Sweden. The catch in Finland was the highest since the mid-1970s, but exploitation rates were decreased on several rivers in Russia in 1991 to provide increased spawning escapement. Rivers in Sweden, along with many in UK, Ireland and France have experienced low flows in 1990 and 1991, and these have had adverse affects on catches.

The marine survival of several monitored European stocks has been low in some recent years, particularly for the 1989 and 1990 smolt year classes. This appears to have been reflected more widely in the poor catches of 1 SW fish in 1990 and both 1 SW and 2SW fish in 1991. Additional information on fisheries in the NorthEast Atlantic is contained in Section 4.

## North America Homewater Fisheries

The Canadian homewater fisheries consist of commercial, recreational, and native food fisheries. There were about 3,300 commercial fishermen licensed to fish for salmon primarily with shore-fast set gillnets. The 1991 fisheries were under quota management with either quotas set for specific salmon fishing areas or for individual fishermen. The total commercial landings in Canada during 1991 were 512 t . Recreational fisheries occurred in all Canadian Atlantic provinces. Anglers were permitted to fish only with artificial flies and were restricted by daily and seasonal retention limits. Retention of salmon of lengths $>63 \mathrm{~cm}$ was permitted only in Quebec and Labrador. Some rivers had specific quotas. In 1991 there were about 282,700 rod days of fishing effort which resulted in a catch of 132 t of salmon. Several native groups were permitted to fish for salmon for food in four provinces (Quebec, New Brunswick, Nova Scotia, and Newfoundland and Labrador). The total harvest in all of these fisheries was 29 t . Commercial fisheries in Canada harvest salmon of USA origin.

The USA homewater fisheries consist only of recreational fisheries in the state of Maine. Anglers were permitted to fish only with artificial flies. There were daily and seasonal retention limits. In 1991, there were 3,157 licensed anglers and a harvest of 238 salmon.

Additional details describing the homewater fisheries are provided in Sections 6.1 and 6.2.

### 5.3 Stock Abundance and Exploitation at West Greenland

To answer the questions posed by NASCO, the Working Group developed and applied a continental run-reconstruction model to estimate abundance of North American salmon at West Greenland (Section 5.3.1). Data necessary to complete this task were available for the fishery years 1983 to 1990 at West Greenland. The "top-down" constraints model in Anon. (1991a) was improved to include an additional constraint related to catches of grilse in Canada during the same year as the fishery in Greenland. Model outputs also were used to derive a range of abundance estimates for North American and European stocks at West Greenland prior to the fishery.

Abundance estimates for North American stocks were then used to define a range of estimates of pre-fishery abundance. A simple model was developed to illustrate the effects of various combinations of catches on the numbers of fish returning to spawn in North America (Section 5.3.2). The effects of these catch combinations were illustrated for varying levels of pre-fishery abundance for 1 SW salmon destined to return as 2 SW spawners.

The implementation of postulated safe catches for various levels of abundance would depend on some pre-season indices of abundance of salmon in the Greenland fishery area. These are considered in Section 5.3.3

### 5.3.1 Determining abundance of North American and European salmon at West Greenland

## Application to North American Stocks

The schematic diagram of the continental run-reconstruction model for North America is given in Figure 5.3.1.1. The purpose of this model is to estimate the population of non-maturing 1 SW salmon (i.e., salmon destined to return as 2 SW spawners) and to estimate fishery area exploitation rates for West Greenland. Prior to commencement of the Canadian and Greenland fisheries, the population of non-maturing 1 SW can be modelled as three distinct groups. The first group will be exploited in the Newfoundland-Labrador commercial fisheries, primarily in SFAs 1-7 and 14a. Because this group is similar in mean size to the maturing 1SW salmon (i.e., grilse returns) it is assumed that the fishery area exploitation rates on non-maturing 1 SW salmon are similar to the maturing component in these Salmon Fishing Areas. The second group will be exploited in West Greenland. Available evidence suggests that this group has limited exchange with the group exploited in Newfoundland-

Labrador (see Anon., 1991a for review). The third group consists of salmon which are unavailable to any fishery prior to return to home waters. Tag recovery information and test fisheries in East Greenland and the Labrador Sea suggest that some fraction of the population is unavailable to any fishery. Following the fisheries, these three groups return about 10 months later to Canadian homewaters where they are exploited as maturing 2SW salmon in commercial fisheries in SFAs $1-13,14 a$ and $14 b$, and in recreational fisheries in rivers.

The primary difficulty in the construction of a runreconstruction model for North American stocks is that the population is simultaneously fished in Canada and West Greenland. Development of fishery area exploitation rates is confounded by an inability to estimate the fraction of the stock present in either fishery. Another difficulty is that the precise age structure of Canadian catches is not estimated annually. Instead, catches are reported as numbers of fish in "small" and "large" size categories. Various sampling programs conducted by Canada however, do allow approximate estimates of the age structure from the known size structure of the catch. The continental run reconstruction model incorporates the uncertainty in estimates of catches, returns to rivers, and known constraints on exploitation rates in Canada to derive a constrained set of exploitation rates which are consistent with the observable data.

The theoretical basis of the model is describe in Anon. (1992c). To facilitate application of the model to North America, key variables are defined in Table 5.3.1.1.

## Constraints on Proportions and Exploitation

The parameters $\mathrm{P}, \mathrm{FU}, \mathrm{Ec}(\mathrm{i})$, and $\mathrm{Eg}(\mathrm{i})$ are all constrained to the $[0,1]$ interval and $0<\mathrm{P}+\mathrm{FU}<1$. Mathematically the constraints can be written
(1) $0<\mathrm{P}<(1-\mathrm{FU})$
(2) $0<\mathrm{FU}<1$
(3) $0<\operatorname{Ec}(\mathbf{i})<1$

$$
\begin{equation*}
0<\operatorname{Eg}(\mathrm{i})<1 \tag{4}
\end{equation*}
$$

The proportion of the 2 SW population returning from Greenland, defined as 1-P-FU, is also constrained to the [0, 1] interval from Eq. 1 and 2. In addition, the parameters $\mathrm{P}, \mathrm{Ec}(\mathbf{i})$, and $\mathrm{Eg}(\mathbf{i})$ are constrained by the estimated number of returns and catches of 2 SW salmon in year ( $i+1$ ), the estimated range of catches of nonmaturing salmon in year $i$, and the observed numbers of returns in the year i. These relationships are illustrated below.

Constraint 1: Range of Non-maturing $1 S W$ catch in Canada in year (i)

Canadian commercial catches are reported as numbers and weight of fish in "Small" and "Large" size categories. Salmon less than 2.7 kg whole weight are graded as small and salmon $>2.7 \mathrm{~kg}$ are graded as large salmon. Small salmon comprise both maturing and non-maturing 1 SW salmon. Large salmon are primarily MSW salmon but some maturing and non-maturing 1SW salmon are also present. Low numbers of previous spawners may also be present in either size category. Let $H_{s}(i)$ and $H_{L}(i)$ be defined as the number of small year i. Raw catch data and return estimates are summarized in Table 5.3.1.2a.

The number of non-maturing 1SW salmon harvested in year $\mathrm{i}\left(\mathrm{Hcl}_{\mathrm{nm}}(\mathrm{i})\right)$ depends on the fraction of 1 SW salmon present in the "large" market category (q) and the fraction of 1 SW salmon that are immature (f_imm).

$$
\begin{equation*}
\operatorname{Hc}_{\mathrm{mm}}(\mathrm{i})=\mathrm{f}_{-} \mathrm{imm} *\left(\mathrm{H}_{s}(\mathrm{i})+\mathrm{q} \mathrm{H}_{\mathrm{L}}(\mathrm{i})\right) . \tag{5}
\end{equation*}
$$

Note that Eq. (5) assumes that all "small" salmon are 1SW fish. Estimates of fractions of non-maturing salmon present in the Labrador-Newfoundland catch were presented in Anon. (1991a). Actual estimates of the immature fraction for the entire fishery have not been developed, but sample estimates suggest a range for f_imm between 0.1 and 0.2 . Similarly, the sea-age composition of the "large" market category is not known with precision. An assumed plausible range of $q$ between 0.1 and 0.3 was examined. Under these assumptions the maximum number of observed non maturing 1 SW salmon in year i would be

$$
\begin{equation*}
\mathrm{Hc} 1_{\mathrm{nm}, \max }(\mathrm{i})=\mathrm{f}_{-} \mathrm{imm}_{\max } *\left(\mathrm{H}_{\mathrm{s}}(\mathrm{i})+\mathrm{q}_{\max } \mathrm{H}_{\mathrm{L}}(\mathrm{i})\right) \tag{6}
\end{equation*}
$$

where $\mathrm{f}_{\mathrm{-}} \mathrm{imm}_{\text {max }}=0.2$ and $\mathrm{q}_{\text {max }}=0.3$.
Similarly, the minimum value of $\mathrm{Hc}_{\mathrm{nm}}(\mathrm{i})$ is

$$
\begin{equation*}
\mathrm{Hc}_{\mathrm{nm}, \min }(\mathrm{i})=\mathrm{f}_{-} \mathrm{imm}_{\min } *\left(\mathrm{H}_{\mathrm{s}}(\mathrm{i})+\mathrm{q}_{\min } \mathrm{H}_{\mathrm{L}}(\mathrm{i})\right) \tag{7}
\end{equation*}
$$

where ${\mathrm{f} \mathrm{imm}_{\text {min }}}=0.1$ and $\mathrm{q}_{\text {min }}=0.1$.
From the run-reconstruction model, Hc1(i) can be expressed as a function of $\mathrm{Ec}(\mathrm{i})$ and P ; call this predicted value ${ }^{\text {hat }} \mathrm{c} 1$ (i). Then

$$
\begin{align*}
& \mathrm{H}^{\mathrm{hat}} \mathrm{c} 1(\mathrm{i})=\mathrm{Ec}(\mathrm{i}) /(1-\mathrm{Ec}(\mathbf{i})) * \mathrm{P} *(\mathrm{R} 2(\mathrm{i}+1) / \mathrm{S} 1+  \tag{8}\\
& \mathrm{Hc} 2(\mathrm{i}+1)) / \mathrm{S} 2
\end{align*}
$$

The feasible range for $\mathrm{Ec}(\mathrm{i})$ and P must yield predicted numbers of nonmaturing 1 SW salmon ( $\mathrm{H}^{\text {hat }} \mathrm{c} 1(\mathrm{i})$ ) between $\mathrm{Hc}_{\mathrm{nm}, \text { min }}(\mathrm{i})$ and $\mathrm{Hc} 1_{\mathrm{mm}, \text { max }}(\mathrm{i})$. Thus the first constraint on the domain of admissible values is

$$
\mathrm{Hc} 1_{\mathrm{nm}, \min }(\mathrm{i})<\mathrm{H}^{\text {hat }} \mathrm{c} 1(\mathrm{i})<\mathrm{Hc}_{\mathrm{nm}, \max }(\mathrm{i})
$$

The estimated number of non-maturing salmon present in the Canadian commercial landings was estimated by restricting the landings only to those Salmon Fishing Areas (SFA) in which non-maturing 1SW salmon (NM_1SW) are believed present in sizable numbers. Therefore, only catches from SFAs 1-7, 14 b were included in the estimates of $\mathrm{H}_{\mathrm{S}}(\mathrm{i})$ and $\mathrm{H}_{\mathrm{L}}(\mathrm{i})$. Minimum predicted values of $H^{\text {hat }} \mathbf{c} 1$ (i) were estimated using $R 2_{\text {min }}(\mathrm{i}+1)$ and $\mathrm{Hc} 2_{\text {min }}(\mathrm{i}+1)$. Maximum predicted values of $\mathrm{H}^{\text {hat }} \mathrm{c} 1(\mathrm{i})$ were estimated using $\mathrm{R} 2_{\max }(\mathrm{i}+1)$ and $\mathrm{Hc} 2_{\max }(\mathrm{i}+1)$ (Table 5.3.1.2b).

By substituting Eq. 8 into 9, it is possible to write Ec(i) as a function of $P$. The upper range of feasible values of $\mathrm{Ec}(\mathrm{i})$ for any P is given by

$$
\begin{equation*}
\mathrm{Ec}(\mathrm{i})=\mathrm{Hc} 1_{\mathrm{nm}, \max }(\mathrm{i}) /\left(\mathrm{Hc} 1_{\mathrm{nm}, \max }(\mathrm{i})+\mathrm{P} * \mathrm{Amin}\right) \tag{10}
\end{equation*}
$$

where $\operatorname{Amin}=\left(\mathrm{R} 2_{\text {min }}(\mathrm{i}+1) / \mathrm{S} 1+\mathrm{Hc} 2_{\min }(\mathrm{i}+1)\right) / \mathrm{S} 2$. Conversely, the lower range of feasible values for $\mathrm{Ec}(\mathrm{i})$ for any P is given by

$$
\begin{equation*}
\mathrm{Ec}(\mathrm{i})=\mathrm{Hc} 1_{\mathrm{nm}, \min }(\mathrm{i}) /\left(\mathrm{Hc} 1_{\mathrm{nm}, \min }(\mathrm{i})+\mathrm{P} * \mathrm{Amax}\right) \tag{11}
\end{equation*}
$$

where $A \max =\left(\mathrm{R} 2_{\max }(\mathrm{i}+1) / \mathrm{S} 1+\mathrm{Hc} 2_{\max }(\mathrm{i}+1)\right) / \mathrm{S} 2$. Equations 10 and 11 define a band of feasible P and $\mathrm{Ec}(\mathrm{i})$ values as shown in Figure 5.3.1.2.

Constraint 2: Range of Estimated 2SW Returns and Catches in Year $(i+1)$

The observed catches of non-maturing 1SW salmon in Canada and Greenland have implications for the total catch of 2 SW salmon and returns to rivers in year $\mathrm{i}+1$. As shown in Anon. (1992c)

$$
\begin{align*}
& (\mathrm{R} 2(\mathrm{i}+1) / \mathrm{S} 1+\mathrm{Hc} 2(\mathrm{i}+1)) *(1-\mathrm{FU})  \tag{12}\\
& =(\mathrm{Tga} 1(\mathrm{i})+\mathrm{Tca} 1(\mathrm{i})) * \mathrm{~S} 2
\end{align*}
$$

The post-fishery populations of 1 SW salmon in Greenland (Tga1(i)) and Canada (Tca1(i)) can be expressed in terms of their exploitation rates as follows:

$$
\begin{align*}
& \operatorname{Tga1}(\mathbf{i})=\operatorname{Hg} 1(\mathbf{i}) / \mathrm{Eg}(\mathrm{i})-\operatorname{Hg} 1(\mathbf{i}), \text { and }  \tag{13}\\
& \operatorname{Tca} 1(\mathbf{i})=\operatorname{Hc} 1(\mathbf{i}) / \operatorname{Ec}(\mathbf{i})-\operatorname{Hc} 1(\mathbf{i}) . \tag{14}
\end{align*}
$$

Substituting Eq. 13 and 14 into Eq. 12 yields,

$$
\begin{align*}
& {[(\mathrm{R} 2(\mathrm{i}+1) / \mathrm{S} 1+\mathrm{Hc} 2(\mathrm{i}+1)) *(1-\mathrm{FU})]^{\text {hat }}=}  \tag{15}\\
& ((\mathrm{Hg} 1(\mathrm{i}) / \mathrm{Eg}(\mathrm{i})-\operatorname{Hg} 1(\mathrm{i}))+(\mathrm{Hc} 1(\mathrm{i}) / \mathrm{Ec}(\mathrm{i})
\end{align*}
$$

$$
\operatorname{Hc} 1(i))) * S 2
$$

Where the "hat" superscript indicates that this is a predicted value, deduced from the values of $\mathrm{Eg}(\mathrm{i})$ and Ec1(i).

Minimum and maximum observed values of the left hand side of Eq. 12 can be derived from the estimated observed range of $2 S W$ runs in year $i+1\left[R 2_{\text {min }}(i+1)\right.$, $\mathrm{R} 2_{\text {max }}(\mathrm{i}+1)$ ] and projections of the range in numbers of 2 SW salmon caught in year $\mathrm{i}+1\left[\mathrm{Hc} 2_{\min }(\mathrm{i}+1)\right.$, $\mathrm{Hc}_{\text {max }}(\mathrm{i}+1)$ ] (see Table 5.3.1.2b). The minimum and maximum numbers of 2 SW salmon caught in year ( $\mathrm{i}+1$ ) are

$$
\begin{align*}
& \mathrm{Hc} 2_{\min }(\mathrm{i}+1)=\mathrm{H}_{\mathrm{L}}(\mathrm{i}+1) *\left(1-\mathrm{q}_{\max }\right), \text { and }  \tag{16}\\
& \mathrm{Hc} 2_{\max }(\mathrm{i}+1)=\mathrm{H}_{\mathrm{L}}(\mathrm{i}+1) *\left(1-\mathrm{q}_{\min }\right), \tag{17}
\end{align*}
$$

respectively. Minimum estimates of $2 S W$ returns to Canadian rivers were estimated by summing the spawning runs reported for some rivers, the total recreational catch and the numbers of salmon counted in traps and at counting fences. The uncertainty in the range of these values led to low and high estimates, $\mathrm{R} 2_{\text {min }}$ and $\mathrm{R} 2_{\text {max }}$, respectively. Thus the range of values predicted in the left hand side of Eq. 15 must lie between estimates of the observed values of $\mathrm{R} 2(\mathrm{i}+1)$ and $\mathrm{Hc} 2(\mathrm{i}+1)$ as illustrated below:

$$
\begin{align*}
& \{\mathrm{A}\}<[(\mathrm{R} 2(\mathrm{i}+1) / \mathrm{S} 1+\mathrm{Hc} 2(\mathrm{i}+1)) *(1-\mathrm{FU})]^{\text {hat }}  \tag{18}\\
& <\{\mathrm{B}\}
\end{align*}
$$

where $\{\mathrm{A}\}=\left[\left(\mathrm{R} 2_{\min }(\mathrm{i}+1)+\mathrm{Hc} 2_{\min }(\mathrm{i}+1)\right)^{*}(1-\mathrm{FU})\right]$, and

$$
\{B\}=\left[\left(R 2_{\max }(i+1)+\mathrm{Hc} 2_{\max }(\mathrm{i}+1)\right) *(1-\mathrm{FU})\right] .
$$

By substituting the right hand side of Eq. 15 into Eq. 18, the exploitation rate in Canada can be written as a function of the exploitation rate in Greenland. The plot of this constraint is shown in Figure 5.3.1.3.

## Constraint 3: Range of Estimated Returns of 1SW Salmon in Year (i)

The final constraint on the derived exploitation rates relates to the exploitation on non-maturing 1SW salmon in year $i$ and subsequent returns to rivers in the same year. If the Ec(i) applies equally to non-maturing and maturing 1SW salmon in SFAs where they are simultaneously fished, then the projected run of 1SW salmon (grilse) is defined as

$$
\begin{equation*}
\mathrm{R}^{\text {hat }}{ }_{1(\mathrm{i})}=\left(\mathrm{Hc} 1_{\text {mat }}(\mathrm{i}) / \mathrm{Ec}(\mathrm{i})-\mathrm{Hc} 1_{\text {mat }}(\mathrm{i})\right) * \mathrm{~S} 1 \tag{19}
\end{equation*}
$$

The minimum and maximum values of Eq. 19 occur when $\mathrm{Hc} 1_{\text {mat }}(\mathrm{i})$ is at its respective minimum and maximum values. Thus

$$
\begin{align*}
& \mathrm{R}^{\text {hat }}{ }_{1_{\min }}(\mathrm{i})=  \tag{20}\\
& \left(\mathrm{Hc} 1_{\text {mal, min }}(\mathrm{i}) / \mathrm{Ec}(\mathrm{i})-\mathrm{Hc} 1_{\text {mal, } \text { min }}(\mathrm{i})\right) * \mathrm{~S} 1
\end{align*}
$$

$$
\begin{align*}
& \mathrm{R}^{\text {hat }{ }_{\max }(\mathrm{i})=}  \tag{21}\\
& \left(\mathrm{Hc} 1_{\operatorname{mat}, \max }(\mathrm{i}) / \mathrm{Ec}(\mathrm{i})-\mathrm{Hc} 1_{\operatorname{ma}, \max }(\mathrm{i})\right) * \mathrm{~S} 1
\end{align*}
$$

where

$$
\begin{align*}
& \mathrm{Hc} 1_{\operatorname{mat}, \max }(\mathrm{i})=  \tag{22}\\
& \left(1-\mathrm{f}_{\mathrm{i}} \mathrm{~mm}_{\min }\right) *\left(\mathrm{H}_{s}(\mathrm{i})+\mathrm{q}_{\max } \mathrm{H}_{\mathrm{L}}(\mathrm{i})\right),
\end{align*}
$$

and

$$
\begin{align*}
& \mathrm{Hc}_{\text {mal, }, \min }(\mathrm{i})=  \tag{23}\\
& \left(1-\mathrm{f}_{2} \mathrm{~mm}_{\max }\right) *\left(\mathrm{H}_{\mathrm{s}}(\mathrm{i})+\mathrm{q}_{\min } \mathrm{H}_{\mathrm{L}}(\mathrm{i})\right),
\end{align*}
$$

where $\mathrm{f}_{-} \mathrm{imm}_{\min }=0.1, \mathrm{f}_{-} \mathrm{imm}_{\text {max }}=0.2, \mathrm{q}_{\text {min }}=0.1$, and $\mathrm{q}_{\text {max }}=0.3$.

The observed range for grilse returns for SFAs 3-7, and 14 a is denoted as $\left[\mathrm{R} 1_{\text {min, obs }}(\mathrm{i})\left\{\right.\right.$ SFA 3-7,14a\}, $\mathrm{R} 1_{\text {max, obs }}(\mathrm{i})$ \{SFA 3-7,14a\}]. Because Eq. 20 and 21 are functions of $\mathrm{Ec}(\mathrm{i})$ only, the minimum and maximum values of exploitation rates on 1 SW salmon in Canada can be expressed as constants by rearranging Eq. 20 and 21 and substituting the minimum and maximum values of $\mathrm{Hc}_{\text {mat }}(\mathrm{i})$ from Eq. 22 and 23. Thus the maximum exploitation rate in Canada, consistent with observed grilse returns and consistent with numbers of maturing 1SW salmon in the catch in SFAs 3-7, and 14a, is given as:

$$
\begin{align*}
& \mathrm{Ec}_{\max }(\mathrm{i})=  \tag{24}\\
& \mathrm{Hc}_{\text {mat, max }}(\mathrm{i}) /\left(\mathrm{R} 1_{\text {min,obs }}(\mathrm{i}) / \mathrm{S} 1+\mathrm{Hc}_{\text {mal, max }}(\mathrm{i})\right)
\end{align*}
$$

Similarly, the minimum exploitation rate consistent with the observed catches and runs is

$$
\begin{align*}
& \mathrm{Ec}_{\text {min }}(\mathrm{i})=  \tag{25}\\
& \mathrm{Hc} 1_{\text {mat, } \min }(\mathrm{i}) /\left(\mathrm{R} 1_{\text {max, obs }}(\mathrm{i}) / \mathrm{S} 1+\mathrm{Hc}_{\text {mat, min }}(\mathrm{i})\right)
\end{align*}
$$

These constraints are illustrated in Figure 5.3.1.2 as the upper and lower horizontal lines.

## Graphical Analysis of Topdown Constraints Model

The constraint equations presented above, represent a system of nonlinear equations which collectively define a feasible range of values for exploitation rates in Greenland and the fraction of the stock available to each fishery. Fortunately, the system of equations can be depicted graphically in the Ec x P and the Ec x Eg planes. The purpose of this section is to illustrate the general properties of the constraints and to derive fishery-area exploitation rates.

To illustrate the estimation of Greenland exploitation rates, the data for the 1988 fishery are presented; the fraction of the population unavailable to any fishery is assumed to be 0.30 . The first step is to plot relationships among the model parameters in the Ec x P plane
(Figure 5.3.1.2). The vertical constraint corresponds to Eq. 1 where $P<1-F U$. The horizontal constraints correspond to Eq. 24 and 25 and are related to the constraints on exploitation in Canada derived from observed returns of maturing 1 SW salmon in year i . Finally, the implications of the estimated numbers of non-maturing salmon in year i catches (i.e., Eq. 10 and 11) are shown in the upper and lower curved lines, respectively (Figure 5.3.1.2).

The relationship between the exploitation rates in year i in Canada and Greenland for observed total catches and returns in year $i+1$ are described in Eq. 15 and 18 and plotted in Figure 5.3.1.3. The upper line corresponds to exploitation rates when $\mathrm{Hc} 1(\mathrm{i})$ is at its maximum value and $\mathrm{R} 2(\mathrm{i}+1)$ and $\mathrm{Hc} 2(\mathrm{i}+1)$ are at their minimum values. Conversely, lower line corresponds to exploitation rates when $\mathrm{Hc} 1(\mathrm{i})$ is at its minimum value and $\mathrm{R} 2(i+1)$ and $\mathrm{Hc} 2(\mathrm{i}+1)$ are at their maximum values. As the constraints on 1SW returns (Eq. 24 and 25) are independent of $\operatorname{Eg}(\mathrm{i})$, they can be superimposed on Figure 5.3.1.3 to constrain exploitation rates in Greenland as shown by the arrows. Thus the maximum feasible range of $\mathrm{Eg}(\mathrm{i})$ $\left\{E g_{\max }(\mathrm{i}), \mathrm{Eg}_{\text {min }}(\mathrm{i})\right\}$ is constrained by the observed range of returns and catches in year $i+1$, the observed range of returns of 1 SW salmon in Canada in year $i$, and the assumed fraction of the population unexploited by any fishery (FU). For the 1988 fishery with $\mathrm{FU}=0.30$, the estimate feasible range of $\mathrm{Eg}(\mathrm{i})$ was $\{0.37,0.55\}$.

The range of exploitation rates in Greenland, $\left\{\mathrm{Eg}_{\text {max }}(\mathrm{i})\right.$, $\left.E g_{\min }(\mathrm{i})\right\}$, can be used to parameterize a set of contour curves for Eg in the Ec x P plane. By definition, the exploitation rates in Canada and Greenland are:

$$
\begin{equation*}
\mathrm{Eg}(\mathrm{i})=\mathrm{Hg}(\mathrm{i}) /((1-\mathrm{P}-\mathrm{FU}) * \mathrm{~A}+\mathrm{Hg}(\mathrm{i})) \tag{26}
\end{equation*}
$$

and

$$
\begin{equation*}
\mathrm{Ec}(\mathrm{i})=\mathrm{Hc} 1_{\mathrm{nm}}(\mathrm{i}) /\left(\mathrm{P} * \mathrm{~A}+\mathrm{Hc}_{\mathrm{nm}}(\mathrm{i})\right), \tag{27}
\end{equation*}
$$

where $\mathrm{A}=(\mathrm{R} 2(\mathrm{i}+1) / \mathrm{S} 1+\mathrm{Hc} 2(\mathrm{i}+1)) / \mathrm{S} 2$. Solving for A in Eq. 26 and substituting into Eq. 27 yields Ec(i) as a function $\mathrm{Eg}(\mathrm{i})$ and P as shown below:

$$
\begin{align*}
& \mathrm{Ec}(\mathrm{i})=\mathrm{Hc} 1_{\mathrm{nm}} /\left((\mathrm{P} /(1-\mathrm{P}-\mathrm{FU}))^{*}\right.  \tag{28}\\
& \left.(\mathrm{Hg}(\mathrm{i}) / \mathrm{Eg}(\mathrm{i})-\mathrm{Hg}(\mathrm{i}))+\mathrm{Hc} 1_{\mathrm{nm}}(\mathrm{i})\right)
\end{align*}
$$

Eq. 28 can be used to generate contour curves for Eg in the Ec $\times \mathrm{P}$ plane by substituting values of Eg between $\mathrm{Eg}_{\text {max }}$ and $\mathrm{Eg}_{\text {min }}$ into Eq. 28. Because $\mathrm{Hc}_{\text {nm }}(\mathrm{i})$ also has minimum and maximum observed values the full set of level curves for Eg must incorporate this measure of uncertainty as well. These relationships are shown in Figures 5.3.1.4 and 5.3.1.5. In each graph, the upper and lower curves correspond to level curves for $\left\{\mathrm{Eg}_{\max }(\mathrm{i}), \mathrm{Eg}_{\min }(\mathrm{i})\right\}$. Note that feasible values of Eg and Ec do not exist for values of P greater than 1-FU.

Figures 5.3.1.4 and 5.3.1.5 can be superimposed on Figure 5.3.1.2 to illustrate the feasible range of exploitation rates within the polygon defined in Figure 5.3.1.2. For the 8 years of data examined, Eq. 28 does not appear to constrain exploitation rates, but it does act to restrict the range of feasible $P$ values.

## Estimation of Exploitation and Abundance

The graphical method of analysis described above was used to estimate feasible ranges of exploitation rates for Canada and Greenland for 1983 to 1990 fishery years (Table 5.3.1.3). The average minimum and maximum exploitation rates for Canada were 57 and $70 \%$ respectively. For Greenland the average minimum and maximum exploitation rates depend on the FU parameter. When the fraction unavailable (FU) was assumed to be 0.05 , the average minimum and maximum exploitation rates in Greenland were 25 and $36 \%$, respectively. Exploitation rates in 1983 and 1984 were particularly low, an observation consistent with the low catches in those years. Exploitation rates between 1985 and 1988 were about twice as high (about 30 to $50 \%$ ); during these years the quota acted to restrict harvests in Greenland. Estimates for 1989 and 1990 are somewhere between the 1983-4 and 1985-8 periods. The 1983-4 and 1989-90 fisheries were unaffected by the quota, suggesting low abundance in the West Greenland area.

When the FU parameter is assumed to 0.3 the estimated range of exploitation rates in Greenland increases over the entire period (Table 5.3.1.3). The same general patterns described above still apply, with low rates in 1983 and 1984, higher rates ranging between 40 to $58 \%$ during the period 1985-88, and intermediate levels in 1989-90. Further examination of environmental factors that may have led to variations in relative abundance of salmon in these fisheries is warranted.

The total estimated abundance of all non-maturing 1SW salmon of North American origin (Figure 5.3.1.6) shows a marked decline since 1986. Estimates in Figure 5.3.1.6 were obtained simply by reconstructing the population for minimum and maximum values of R2(i+1 ), $\mathrm{Hc} 2(\mathrm{i}+1), \mathrm{Hc}_{\mathrm{nm}}(\mathrm{i})$ and $\mathrm{Hg}(\mathrm{i})$ (see Table 5.3.1.2b, and Eq. 1 in Section 5.3.2). Thus, these estimates represent the entire extant stock. While the data do not indicate abundance by fishing region, the estimates illustrate an over two-fold range of pre-fishery abundance in an 8 year period.

The total estimated abundance of all salmon in the West Greenland area can be estimated by dividing the total catch by the minimum and maximum values of exploitation rates $\left\{\mathrm{Eg}_{\min }(\mathrm{i}), \mathrm{Eg}_{\max }(\mathrm{i})\right)$. The derived range of abundance estimates (Figure 5.3.1.7) suggest a general downward trend since 1985, regardless of whether $\mathrm{FU}=0.05$ or 0.30 . Peak abundance in 1985 probably
ranged from 800,000 to 1 million non-maturing salmon of all sea ages (mostly 1SW). Trends for European and North American stocks appear to be more erratic, but both stock complexes exhibit very low abundance in 1989 and 1990 (Table 5.3.1.3).

Derived ranges of $P$, i.e., the fraction of the stock in Canada, show little trend over time (Figure 5.3.1.8). A relatively small proportion of the 2 SW returns appear to return from Canadian fisheries on non-maturing 1SW salmon. Model results suggest that most of the 2SW returns come either from Greenland or are unavailable to any fishery. The relative proportions returning from Canada ( P ), Greenland (1-P-FU) are contingent on the assumed fraction of the stock unavailable (FU). Model results can suggest upper bounds on FU but actual estimates may be impossible to obtain. Analysis of marine environmental conditions, however, may allow imputation of feasible values. In view of the important questions raised by these model results, the Working Group recommended further assessment. In particular, an examination of potential causative environmental factors is warranted.

## Discussion

The derived estimates of fishery area exploitation rates apply collectively to most of the North American stocks that frequent the West Greenland area. To the extent that different stocks have different migration patterns, the period of residence within the fishery would determine the actual rate of exploitation on that stock. Fish that reside within the fishery for longer periods would have greater exploitation rates.

This modelling approach could be applied to specific stocks when data are available. The modellng approach has been applied to all Canadian stocks which have a significant proportion of MSW spawners. For these data, the derived exploitation rates apply to the entire group of stocks and therefore represent an average rate for that fraction (i.e., 1-P-FU) of the population available to Greenland. The input data could be further disaggregated to incorporate stock complexes, such as northern and southern Canada rivers. Further disaggregation for specific stocks would be possible when specific harvest data were available (e. g., Maine-origin salmon). The important point is that the scope of inferences regarding fishery area exploitation rates would necessarily be reduced. Application to northern and southern Canadian stocks could however, highlight important differences in exploitation rates relevant to improving future management decisions. Potential applications of this refinement are highlighted in Section 5.3.2.

In light of the potential application of this approach to interpret historical patterns of abundance and exploitation rates, the Working Group recommends that total return
and catch data be assembled for years (i.e., fishery years) prior to 1983.

## Application to European Stocks

The Working Group considered ways to apply similar models to the European stocks exploited at West Greenland. Because of the nature of the fisheries, the constraints model cannot be applied directly. The Working Group therefore considered a preliminary estimate of the abundance of non-maturing 1 SW salmon in the sea at the time of the West Greenland fishery based upon a runreconstruction approach. Catches of 2SW salmon in homewater fisheries were used to estimate the numbers of 2SW fish returning to each country. These were then used to estimate the numbers that would have been alive in the previous year.

Maximum and minimum estimates of the numbers of non-maturing 1SW salmon alive at the time of the West Greenland fishery for each European country [T1g(max) and $\mathrm{T} 1 \mathrm{~g}(\mathrm{~min})$ ] were calculated using the following formulae:

$$
\begin{aligned}
& \mathrm{T} 1 \mathrm{~g}(\max )= \\
& \mathrm{Hm} / \mathrm{q} /(1-\mathrm{R}) / \operatorname{Exp}(-\mathrm{Mt}) / \mathrm{E}(\max ) \\
& \mathrm{T} 1 \mathrm{~g}(\min )=\mathrm{Hm} / \mathrm{q} /(1-\mathrm{R}) / \operatorname{Exp}(-\mathrm{Mt}) / \mathrm{E}(\min )
\end{aligned}
$$

where:
$\mathrm{Hm}=$ harvest of MSW fish in homewaters in year $\mathrm{i}+1$;
$\mathrm{R}=$ average non-reporting rate for all European countries;
$\mathrm{q}=$ proportion of Hm that comprises 2SW fish;
$\mathrm{M}=$ natural monthly mortality rate for 1 SW salmon in the sea;
$\mathrm{t}=$ months from West Greenland fishery to homewater fishery;
$\mathrm{E}(\max )=$ maximum likely mean exploitation rate on 2SW fish in homewaters;
$\mathrm{E}(\mathrm{min})=$ minimum likely mean exploitation rate on 2SW fish in homewaters;

The assessment was carried out for 1990 catches in homewaters. This gave estimates of the numbers nonmaturing 1SW European salmon alive before the 1989 West Greenland fishery of between 915,000 and 1,242,000 salmon (Table 5.3.1.4). Table 5.3.1.2b gives a range of estimates of the total abundance of salmon in the West Greenland fishery area in 1989 from 317,351 to 618,000 . Using the proportion of European salmon estimated to be in the catch that year ( $44 \%$ ), the numbers of European fish in the area is estimated to be between 139,635 and 271,920 . This therefore suggests that between $11 \%$ and $30 \%$ of all the non-maturing 1 SW European salmon were in the fishery area in 1989.

It is important to note, however, that there is a marked difference in the proportions of the stocks from the northern and southern countries that go to the West Greenland area. The Working Group recommended that further analyses be carried out to examine the different contributions of stocks from different areas to the West Greenland fishery.

### 5.3.2 Modelling interactive effects between abundance and exploitation rates at West Greenland in relation to achievement of North American spawning targets

The pre-fishery abundance of North American nonmaturing 1 SW salmon (i.e. 2 SW spawners) can be estimated by applying the run-reconstruction model to total returns and catches for Canada. The term "prefishery" refers to the population size prior to commencement of the Canadian and Greenland fisheries. This population estimate does not include individuals of the same smolt class which mature and return as 1SW spawners. The estimated pre-fishery abundance of nonmaturing 1SW salmon in fishery year (i) is defined herein as N 1 (i), and estimated as:

$$
\begin{align*}
& \mathrm{N} 1(\mathrm{i})=(\mathrm{R} 2(\mathrm{i}+1) / \mathrm{S} 1+\mathrm{C} 2(\mathrm{i}+1)) / \mathrm{S} 2+\mathrm{C} 1(\mathrm{i})  \tag{1}\\
& +\mathrm{G} 1(\mathrm{i})
\end{align*}
$$

where
G1(i) Catch of 1SW salmon in Greenland in year i
C1(i) Catch of non-maturing 1SW salmon in Canada in year i

R2(i+1) 2SW Spawners to Canadian rivers in year $\mathrm{i}+1$

C2(i) Total Catches of 2SW salmon in Canada in year i+1

S1 Survival rate between the fishery on $\mathrm{C} 2(\mathrm{i}+1)$ and returns R 2 (i) to the river

S2 Survival rate between the midpoint of the C 1 (i) and G1(i) fisheries and the $\mathrm{C} 2(\mathrm{i}+1)$ fishery.

Estimates of total pre-fishery abundance of 1SW salmon destined to return as 2 SW spawners are presented in Figure 5.3.1.6.. Data for 1983 to 1990 are presented in Table 5.3.1.2b.

In the analyses which follow, S1 represents the application of a constant annual mortality rate ( $M=0.12$ ) over a period of 2 months. The parameter S 2 corresponds to an annual M of 0.12 applied over a period of 10 months. For notational simplicity, the subscripts on G1, C1, C2, R 2 , and N 1 will be dropped in the following paragraphs.

Equation 1 can be rearranged to express R 2 as a function of C1, G1, and C2 for varying levels of N1. To illustrate potential utility of the approach, various combinations of catches ( $\mathrm{G} 1, \mathrm{C} 1$, and C 2 ) on estimated numbers of spawners were computed for various levels of N1. Results illustrate that a wide variety of catches would allow equivalent numbers of spawners to return.

The target number of spawners necessary to achieve conservation objectives can be called R2_target. The Working Group considered a provisional estimate of R2_target of about 175,000 which represents the sum of target spawning requirements for all Canadian rivers. At low levels of pre-fishery non-maturing 1SW abundance (N1 = 200,000; Figure 5.3.2.1) there would be insufficient numbers of spawners (R2) to allow harvest in either Canada or West Greenland. At moderate ( $\mathrm{N} 1=400,000$ ) and higher ( $\mathrm{N} 1=600,000$ ) levels of abundance (Figures 5.3.2.2-3) a range of catch allocations among fisheries (C1 vs G1) or years ( $\mathrm{G} 1, \mathrm{C} 1$ vs C 2 ) would permit sufficient numbers of spawners within "safe biological limits", provided that targets for component stocks met. Based on observed projections since 1983 (Figure 5.3.1.6), a reasonable range of N 1 values is 200,000 to 600,000 salmon.

The Working Group identified several problems with using abundance and exploitation information to provide management advice, especially in relation to spawning targets. Although the combined target spawning requirements for Canadian rivers is probably between 150,000 and $200,0002 \mathrm{SW}$ salmon, meaningful catch advice to provide sufficient spawning escapement for individual stocks is not readily available due to the varying proportions of stocks contributing to the fisheries. Previously, the Working Group (Anon., 1982, 1984) advised that "it is not possible at the present time to estimate and advise on a single TAC which would maintain homewater stocks and safeguard the spawning within safe biological limits". Anon. (1984) further advised that regulation by a single TAC would not seem to be a practical method to adequately ensure spawning escapement within safe biological limits for stocks which are, in part, harvested in mixed stock fisheries.

The present Working Group, however, noted that if current catches are adversely affecting the total stock, then reductions in catches would benefit the population as a whole. Benefits to specific stocks however, could not be predicted. Development of a TAC that would reduce catches when stocks are low would provide means of indicating when catch reductions are biologically justified. Present methods for setting catch levels irrespective of population size pose an even greater risk to the total population during periods of low stock abundance. Future management advice could be improved as additional information on particular stocks became available.

The Working Group considered two approaches for improving catch advice. Spawning targets could be improved by taking known individual river spawning targets and scaling these up regionally to identify a minimum overall North American target. Another alternative would be to group North American 2SWproducing stocks into "stock complexes" based on river age distributions and evaluate their contribution to catches in Canada and Greenland. The Working Group recommended that both approaches to improvements be evaluated and reviewed at its next meeting.

A significant problem with providing catch advice based on average abundance levels is that it does not account for differences among stocks in productivity or their contributions to fisheries, and hence ability to sustain particular levels of exploitation. This generates a risk of driving already weak stocks further downwards, while under-exploiting stocks of higher productivity.

The general principles relating the productivity of stocks to their ability to withstand exploitation are illustrated in Figure 5.3.2.4. Salmon stocks differ widely in their natural productivity. In addition, human influences on salmon rearing habitat may have reduced the productivity of some systems. In Figure 5.3.2.4, the vertical axis shows the number of returning adult fish that would be produced for each spawner in the previous generation, in the absence of exploitation. This value will vary between stocks and thus determines the position of the stock on the horizontal axis.

Stocks lying at point Eo are just self-sustaining in that each spawner will only produce one returning adult fish, leaving no extras available to be harvested. To the left of this point lie stocks that cannot produce enough spawners and are thus not self sustaining. Such stocks could not exist naturally, however some have been reduced to this condition by factors damaging the river habitat; these stocks can now only be maintained by stocking. To the right are rivers that are producing more fish each generation than are required for adequate spawning and can therefore tolerate exploitation without declining. Most salmon rivers are thought to fall in this latter group.

The effect of exploitation is to reduce the number of adults returning to spawn. Thus a fixed level of exploitation on a wide range of stocks is illustrated by the dotted line in the figure. This has the effect of putting a greater proportion of stocks beyond the point at which they are under threat or are not self sustaining.

The real position will clearly be more complex than this for a number of reasons. Among the most important of these is that the productivity of a particular system is likely to alter between years. In addition, mixed stock fisheries will normally exploit different stocks at different
levels.

Unless exploitation occurs only in rivers, any mixed stock fishery will increase the risk of over exploiting individual stocks. The relevant scientific issue, in the context of current management practices on mixed stock fisheries, is to reduce the risk to less productive stocks and stocks that contribute greater relative proportions to the fisheries. Without an extensive tagging program, river-specific stocks cannot be identified in the mixed stock fishery at Greenland or elsewhere. Thus, any level of exploitation in the mixed stock fishery potentially poses problems for some stocks. Figures 5.3.2.1-3 illustrate the tradeoffs that are necessary to achieve numbers of spawners in home waters equal to the combined spawning requirements of all stocks. These tradeoffs apply in general, but not to specific stocks.

Identifying specific times and areas where vulnerable stocks are present in the fishery could provide protection, but analyses to date have not revealed predictable patterns for any North American stock. Alternatively, excessive exploitation on the 1SW component, when it occurs, can be compensated by reductions in catches on the maturing 2 SW returns. The feasibility of this approach depends on the magnitude of the exploitation on the 1 SW component. In addition, feasibility also depends on non-biological factors related to desired allocation of catches among fisheries.

Future refinement of this approach would include assessment of the relative risks associated with errors in the prediction of N1 and application of the model to groups of stocks, e.g., northern and southern Canadian stocks.

### 5.3.3 Indices of abundance at West Greenland

The Working Group examined information from Canada which may provide a pre-season index of abundance of North American fish at West Greenland. Information on 1SW returns to several Canadian rivers, catches of small salmon in Canadian waters and survival rates of hatchery 1SW salmon to some Canadian rivers were examined in relation to part or whole season measures of abundance in the Greenland fishery. Among several significant relationships, (Anon. 1992a), the predictor judged to have the greatest management potential was the count of "small" salmon in the Millbank trap on the Miramichi River. If the statistical prediction error of this relationship were shown to be sufficiently small, then 1SW numbers in this trap could provide a pre-season index of abundance of 1 SW Canadian salmon in the Greenland fishery. The significance of the relationship ( $\mathrm{P}=0.001$ ) suggests that variations in the magnitude of post-smolt survival exceeds the variability in proportion of fish maturing as 1 SW salmon. Thus, variations induced in the regression line by interannual fluctuations in maturity
and variations in smolt production are not sufficient to obscure the underlying relationship. This means that in the Miramichi stock, the overall rate of return for a smolt class is dominated by post-smolt mortality rather than variations in maturation rates.

The Working Group considered information from catches of small salmon in southern Labrador (SFAs $2 \& 3$ ) that could be used to predict current year abundance at West Greenland. The relationship of numbers of North American river age 4 and older caught in the first two weeks of fishing at West Greenland on catch in number of small salmon in Labrador was significant at less than the $5 \%$ level of significance $\left(\mathrm{R}^{2}=0.66, \mathrm{~F}=17.8, \mathrm{P}=0.0-\right.$ 02). In order for this relationship to be of use to management as an index of abundance of salmon at Greenland, a data series of catches of small salmon in Labrador up to a specific date would have to be developed.

Another concern with the use of pre-season indices is that they apply only when management measures have not altered the basic underlying relationship with measures of abundance in Greenland. For example, closure of a major fishery could confound the interpretation of data obtained from remaining fisheries.

Factors influencing potential indices of abundance of European 1SW salmon at West Greenland were evaluated by the Working Group. Potential indices were thought to fall into two main groups; those related to the production of juveniles or smolts; and those based on estimates of the abundance of 1 SW salmon. Run-reconstruction modelling on monitored stocks was used to examine the relative variation in post-smolt survival and the proportion of salmon maturing at 1SW. This demonstrated that among European stocks variability in the proportion of a stock maturing at 1 SW was low in comparison to the variation in post-smolt survival (Table 5.3.3.1).

These results suggest, but do not confirm, that returns of maturing 1SW salmon to some European rivers could serve as indices of the abundance of non-maturing 1SW in the sea at the time of the West Greenland fishery. It was felt that many European stocks would be unlikely to provide a pre-season index of abundance at West Greenland due to the 1 SW returns being spread over the middle and latter parts of the year and being very variable. However, the Working Group recommended that data from fisheries, river counts and traps be examined further to evaluate this possibility.

An alternative approach, involving use of early season daily catches in the Greenland fishery as an index of abundance, was considered by the Working Group. If catches during the early part of the season are correlated to abundance, then the rate at which catch accumulates might be used to indicate potential abundance in the same
season, and thus be used for catch advice. However, it was thought that as catches are dependant on several factors in addition to abundance, catch alone might not be a suitable index of current season abundance. The Working Group recommended that further investigation of past patterns of abundance and distribution of early season catches be carried out. Preliminary assessment of this approach (Anon., 1992a) for 4 years of daily catch data (1987-1990) suggested that within season forecasts of abundance were not sufficiently accurate to provide meaningful advice. However, this assessment examined daily catch patterns for each year individually. A more rigorous approach, using all the data to develop predictive relations, would be useful. Application of the method of Noakes (1989) is recommended. Preparation and analysis of all available daily catch information, prior to the next meeting of the Working Group, is recommended.

### 5.3.4 Exploitation of Maine-origin (USA) salmon

## Extant Exploitation Rates

In (Anon., 1990a), a model to calculate extant exploitation rates of 1SW and 2SW Maine-origin salmon was presented. These calculations were updated by including the 1990 Greenland and Newfoundland catches and 1991 homewater returns (Tables 5.3.4.1-5.3.4.3). The extant exploitation rates for 1SW Maine salmon in 1990 were lower than in the previous year but still higher than the long-term average. The extant exploitation rates for 2SW salmon in 1990 were higher than the average for the time series.

## Fishery area exploitation on 1SW Maine origin salmon at Newfoundland and West Greenland

Fisheries for non-maturing 1SW salmon of North American origin salmon occur simultaneously in West Greenland and Canada. Exploitation rates in these fisheries depend on what proportion of the extant stock is vulnerable to each fishery. Assuming that the population of 1SW Maine-origin salmon is available to only the Newfoundland and West Greenland fisheries, estimates of exploitation rates are presented in Table 5.3.4.4. A monthly natural mortality rate of 0.01 was used in all cases. Two assumed levels of reporting are included; where the Carlin adjustment of 1.0 is used, the reporting rate is that which has been used in Section 5.1.3, and where the Carlin adjustment of 2.0 is used, the reporting rates are halved. Varying levels of $P$ between 0.1 and 0.9 (the proportion of stock migrating from Canada) are evaluated.

The values calculated for 1990 show exploitation in Canada and Greenland are unchanged compared to the previous year. The effects of different reporting rates of Carlin tags and different proportions of the stock popula-
tion available to each fishery, are presented in Figure 5.3.4.1 and these indicate the possible range of fishery area exploitation in 1987-1990.

Estimates of exploitation rates for Maine stocks in Canada and Greenland are generally higher than those estimated by the continental run reconstruction model presented in Section 5.3.1. Those estimates of fishery area exploitation rates are based on the aggregate behaviour of many hundreds of stocks. Maine stocks are near the southern boundary of Atlantic salmon habitat and are likely have different migration routes than the major Canadian stocks. As exploitation in fisheries are dependent upon fishing effort and residency of fish with the area, differences among stocks are to be expected. Comparisons of tag recoveries and harvest estimates between Maine stocks with other North American stocks could provide useful information on different vulnerabilities of individual stocks to fisheries.

### 5.4 Advice on Catch Levels at West Greenland

The Working Group was asked to propose and evaluate methods to estimate possible catch levels based upon maintaining adequate spawning biomass. The problems of estimating the total allowable catch (TAC) for salmon have been examined by the Working Group (Anon., 1982, 1984, 1986, 1988a). The general concerns about the difficulties of applying a TAC are expressed in Section 5.3.2. Although advances have been made in our understanding of population dynamics of Atlantic salmon and the exploitation occurring in the fisheries, the concerns about the implications of application of TACs to mixed stock fisheries are still relevant. The corollary of previous concerns is that reductions in catches, provided via an annually adjusted TAC would, in principle, reduce mortality on the population as a whole. Benefits that might accrue to particular stocks would be difficult to demonstrate, in the same way that detriments to individual stocks are difficult to identify.

The Working Group considered how the predictive measures of abundance, outlined in Section 5.3, could be implemented to give annual catch advice. The aim of advice would be to limit catch to a level that would facilitate achieving overall spawning escapement equivalent to the sum of spawning targets in individual North American and European rivers (when the latter have been defined). To achieve the desired level of exploitation, for a given level of predicted abundance, either a TAC could be fixed or some form of effort limitation introduced.

Although a TAC would limit total catch, it will concentrate the fishing mortality on the stocks in the fishing area up to the time when the various boat quotas are filled. If different stocks enter and leave the fishery area at different times, this could increase the risk of certain
stocks being more heavily exploited than others.
Effort limitation would, in theory, provide a greater range of options for management, such as season length restrictions, regulating number of boats or licenses or close periods in the fishery. However, it was felt that the diversity of boat types and sizes and their large numbers this would make effort limitation difficult in practice, particularly because no reliable data exist on the relationship between effort and exploitation in the fishery.

The methodology employed in Section 5.3 .3 simply defines the tradeoffs in catches of non-maturing 1SW salmon in Canada and Greenland and 2SW catches in Canada in the following year. In particular, it defines a set of feasible combinations that may ensure that an overall spawning target is met. The advice for any given year is dependent on obtaining a reliable predictor of total non-maturing 1SW abundance for North American stocks. Since pre-fishery abundance for year i is the sum of the catches in year $i$ and catches plus returns in year $i+1$, the advice for year $i+1$ fisheries ( $2 S W$ ) could be improved by updating the prediction conditioned on the 1SW catches in year i. For the 1983-1990 data the regression between total 2 SW returns plus 2SW catches in year and total 1 SW catches had a coefficient of determination of 0.76 . More importantly, the standard error of the prediction was 25,000 fish. Hence, management corrections for $2 S W$ catches may be possible.

In contrast, prediction of pre-fishery abundance of 1SW salmon destined to return as 2 SW salmon is much more difficult, as described in previous sections. One possibility would be to use simple trend analysis of the abundance data in Figure 5.3.1.6 to project future abundance. Such predictions could have wide prediction intervals and it would be important to proceed cautiously by using the lower range of predicted abundance levels for management decisions. Further analysis of the error structure of the N1 abundance estimates might provide a means of imputing error bounds on the projections. In turn, these error bounds could be incorporated into the catch advice and expressed in terms of the likelihood of achieving spawning targets. Some portion of the stock will receive excess numbers of spawners while others will be under seeded. However, the above approach could provide improved safeguards, particularly if applied to smaller geographical groupings of stocks (stock complexes) as suggested by the Working Group in 1988 (Anon. 1988a).

### 5.5 By-catches in the Greenland Salmon Drift-Net Fishery

By-catch information for the West Greenland salmon drift-net fishery is not routinely recorded. The only information available was collected during research investigations.

Information on by-catch at West Greenland comes from three sources:

1) Canadian research cruises in Greenland waters during the fishing seasons over the period 1969 to 1982 collected information about by-catches.
2) In 1972 from August to October when the Joint ICES/ICNAF Salmon Tagging Experiment at West Greenland 1972 took place, the by-catch information from four research vessels and from eight commercial vessels was obtained (Christensen and Lear, 1977). On the basis of catch-per-unit-effort of by-catches by these commercial vessels, estimates have been derived on numbers of individuals of various species captured by non-Greenlandic vessels fishing for salmon in 1972. The most numerous by-catches of seabirds were murres (Uria lomvia), dovekie (Plautus alle), greater shearwater (Puffinus gravis), black guillemot (Cepphus grylle) and Atlantic Puffin (Fratercula arctica). Cod (Gadus morhua) was the only fish species taken in significant quantities as by-catches. Harbour porpoise (Phocoena phocoena) and seals were the only marine mammals taken frequently.
3) Piatt and Reddin (1984) provided an analysis of seabird by catches in the salmon drift-nets during the fishing season in the beginning of the 1980s.

The by-catch rates derived from the research vessels in the 1970s and in 1980s as reported by Christensen and Lear (1977) and Piatt and Reddin (1984) are not applicable to the present fishery. The fishing pattern in the Greenlandic salmon drift-nets fishery has changed considerably since the 1970s. There are no longer large drift-net vessels; therefore, the salmon fishing now takes place mostly close to the shore or in the fjords. Since the beginning of the 1980s, the fishery has operated in more southerly areas than before.

### 5.6 Adequacy of Sampling Program at West Greenland

The sampling program at Greenland started in its present form in 1985. The purpose is to sample the fishery as adequately as possible both geographically and temporally. Also, the sample sizes have to be of sufficient size to permit establishment of a reliable data base. As the sampling program is planned in detail before the fishing season, it must draw on the experiences from earlier years. Furthermore, the sampling program is limited to its present size by the amount of money allocated to it by the governments concerned. At present the sampling is carried out by three sampling teams, one in each of three NAFO divisions, during two-four week periods in the
beginning of the fishing season.
The adequacy of the program was discerned by examining the percent of the catch taken during the sampling program and the percent of catch sampled for fork lengths and CWTs (Table 5.6.1). The percent of the catch sampled from 1985-91 has ranged from $25.6 \%$ to $92.1 \%$. A particularly low coverage occurred in 1990 when the catch was low and did not peak early in the season.

The Working Group discussed several ways of improving the sampling program. Scenario 1 in the text table below shows the impact of adding one or two weeks to the present program of sampling in three NAFO Divisions. Scenario 2 was achieved by dropping a NAFO Division and adding a team to another site after the first teams have left.

| Percent of catch in sampling period |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Year | Present | Scenario 1 |  | Scenario 2 |
|  |  | Present |  |  |
|  |  |  |  |  |
|  |  |  |  |  |
|  |  |  |  |  |
|  |  |  |  |  |
|  |  | wk |  |  |
| 1985 | 80.0 | 88.8 | 93.9 | 97.9 |
| 1986 | 92.1 | 94.5 | 97.1 | 99.2 |
| 1988 | 79.3 | 84.7 | 89.2 | 89.2 |
| 1989 | 74.1 | 74.8 | 96.3 | 96.3 |
| 1990 | 25.6 | 81.9 | 86.8 | 98.2 |
| 1991 | 51.9 | 61.7 | 58.6 | 86.9 |

If the objective was to sample during the period when $80 \%$ of the catch was taken then it could only be achieved by Scenario 2. Unfortunately, this would result in poorer coverage over NAFO Divisions. An analysis of temporal and spatial differences in proportions of salmon of North American origin occurring within each fishing season since 1969 was presented. This analysis showed very few indications of temporal changes, while on the other hand significant results occurred in more than $50 \%$ of all analyses concerning spatial variations. This would suggest that scenario 2 is inferior to scenario 1 as it has poorer coverage over NAFO Divisions. It should also be pointed out that it is not always possible to sample small catches at the tail end of the fishery due to the logistics of plant operation.

The sampling program may influence the reporting rate of external tags. It was of particular concern if the reporting rates within and outside the sampling period were different (Table 5.6.2). There was no indication that the reporting rate was always higher within or outside the sampling period. In 1986 and 1988, the
reporting rates outside the sampling periods were significantly higher than within the sampling periods. This was contrary to expectations and no good explanations could be found. However, in 1986 the sampling teams encouraged reporting of Carlin tags and this may have contributed to an increased reporting rate at the end and after the sampling season. Considering the low number of tags it seems likely that any non-random distribution of tagged salmon may have considerable impact on the results. Furthermore, the date when the landings are recorded does not always coincide with the actual catch date.

The proportion of the catch examined for CWTs has varied in the range $4.46-13.26 \%$ of total catch, while the absolute numbers examined varied between 6,41030,360 . The lowest numbers occurred in 1990 and 1991. The Working Group considered this to be of potential concern as low numbers of CWTs recovered will contribute to a decrease in the current database regarding the number of salmon with known origin. It may therefore be worthwhile to try to increase the number of salmon examined for CWTs.

In all other aspects the present sampling program was considered adequate, as long as the catch peaks early in the fishing season.

### 5.7 Effectiveness of Management Measures

Over the past 16 years, the fishery at West Greenland has been managed by a TAC which has ranged from 840 $t$ to 1265 t (Table 5.1.1). In seven years the full quota has not been taken and so the management measures have not restricted the effect of the fishery on stocks returning to homewaters. For the remaining years, it has not been possible to estimate the catch that would have been taken in the absence of a quota control. As a result, the effect of the quota on stocks returning to homewaters is not known.

In the past eight years the quota has been fairly stable (between 840 and 935 t ). Over the same period, exploitation is estimated to have remained at a relatively constant level of approximately $30-45 \%$ (Table 5.3.1.3). As a result, the catch has varied with the abundance of stocks in the area. Data from homewater fisheries (Section 3) suggest that many stocks have been in decline in this period and this is particularly evident in the MSW component of the stocks.

## 6 FISHERIES RELEVANT TO THE NORTH AMERICAN COMMISSION

### 6.1 Description of Fisheries in Canada

The following were new management measures for commercial fisheries in 1991:

In 1991, quotas for the Newfoundland commercial salmon fishery were lower by the following amounts in these Salmon Fishing Areas (SFAs, Figure 3.3.1.1) of Newfoundland: SFA 3 ( 35 t ), SFA 4 ( 22 t ), and SFA 13 ( 10 t ). The quotas in 1990 and 1991 are shown in Table 6.1.1. Salmon Fishing Area 1 had an allowance of 80 t , the same as 1990; an allowance is an estimate of expected catch and not a limitation on allowable harvest. In other SFAs, quotas remained as in 1990. Monitoring of the quotas was conducted by Fisheries Officers who were in contact with buyers and fishermen on a weekly or daily basis.
2) In the Québec commercial fishery, the quota in Q7 was reduced by $34 \%$ (from 2755 to 1809 fish), commensurate with a reduction in a number of licenses under a buy-back program. In Q8 and Q9, the quota and fishing seasons remained essentially the same as they were in 1990.

The following were new management measures for recreational fisheries in 1991:

1) The seasonal bag limit for the recreational fishery of Newfoundland-Labrador was reduced from 15 to 10 fish. For conservation reasons, most rivers in SFAs 22 and 23 (Inner Bay of Fundy) were not opened to recreational fishing.

A more detailed description of other aspects of the commercial fisheries was provided in Anon. (1985) and updated annually by the Working Group.

The total salmon landings for Canada in 1991 were 679 t (Table 2.1.2); this is the lowest recorded landing in the 1960-91 data set. The landings of small salmon ( 318 t ) in 1991 were $25 \%$ below the 1990 landings ( 425 t ) and $51 \%$ below the previous 5 -year mean ( 653 t ). The landings ( 361 t) of large salmon in 1991 were $26 \%$ below the landings of large salmon in 1990 ( 486 t ) and $47 \%$ below the previous 5 -year mean ( 688 t ). Of the total Canadian landings by weight, $25 \%$ were in Québec, $68 \%$ in Newfoundland and Labrador, and $8 \%$ in the Maritime Provinces. The recreational fisheries harvested $20 \%$, commercial fisheries $75 \%$, and the native food fisheries $4 \%$ of the total landings by weight (Table 6.1.2). The decline in total commercial landings from 1596 t in 1987 to 512 t in 1991 was spread over all Salmon Fishing Areas of Newfoundland and Labrador and the Québec North Shore (Table 6.1.1). Landings in Newfoundland and Labrador of 434 t in 1991 were the lowest recorded for the data sets (1910-1970) provided by May and Lear (1971) and 1971-91 (Table 6.1.3). Landings in 1990 were the second lowest in the data sets. A description of the commercial, recreational, and native fisheries, in 1991, in Atlantic Canada are provided in the North American Study Group Report (Anon. 1992a).

The landings in the commercial (in weight) and in the recreational fisheries (in numbers) in 1991 by SFA and comparisons to the mean landings, 1984-89, are provided in Tables 6.1.4, and 6.1.5, respectively, and in Figures 3.3.1.2. Historical commercial and recreational landings are presented in Figure 6.1.1. Analyses of the effects of the 1991 quotas are contained in Section 6.4.

The commercial landings of small and large salmon, by weight, declined from the mean landings 1984-89 throughout Newfoundland and Labrador SFA 1 to 14 (Table 6.1.4 and Figure 3.3.1.2). The early closure in SFAs 3, 5, 6, 7, 8, 9, 10, 11 and 13 contributed to the decrease in landings. However, low abundance of salmon also appears to be a contributing factor, particularly in SFAs 1, 2 and 4. Severe ice conditions along the coasts of NE Newfoundland and Labrador interfered with fishing activities which may also have been a contributing factor to low landings. In these areas catches of large salmon were about three weeks later in 1991 than average while small salmon catches were approximately one and a half weeks later. This apparent delay in migration timing may also be related to the unusually cold ocean water temperatures.

The landings of small and large salmon in Q 7 and Q 8 , the Québec North Shore, decreased from the mean landings 1986-1990, but remained steady in Q9 (Table 6.1.4 and Figure 3.3.1.2). The declines in large salmon catch in Q7 ( $25 \%$ ) may be related to the reductions in quotas and removal of fishing licenses.

The recreational landings of large salmon declined in all Salmon Fishing Areas except Q2, Q3, Q5, and Q8 (Table 6.1.5 and Figure 3.3.1.2). Similarly, declines were experienced in small salmon in all SFAs, except SFA 3, 18, Q2, Q3, Q5, Q6, and Q9 (Table 6.1.5 and Figure 3.3.1.2).

### 6.1.1 Composition and origin of the catch, 1991

Salmon of Canadian and USA origin were captured in the fisheries of Newfoundland and Labrador, in 1991, as evidenced by recaptures of tagged 1 SW salmon from both countries. Salmon with coded wire tags (CWT) were recovered during sampling at 12 commercial fishing ports in SFAs 2, 3, and 4. Samplers scanned over 19,000 fish (about $21 \%$ of the total catch in these SFAs) for the presence of adipose fin clips and CWTs. Of this sample, 159 fish had adipose fin clips and 120 fish had CWTs (Table 6.1.1.1). As in 1990, the highest percentage of CWT-positive salmon were observed at Harbour Deep (SFA 3) with an occurrence rate ( $2.5 \%$ ) nearly twice as great as any other fishing port. Overall, the presence of CWT-positive salmon was 2.8 -fold greater in 1991 than in 1990.

Ninety-seven percent of the 120 CWTs recovered were
from USA rivers. In contrast to previous years, most tag recoveries were from the Connecticut River (74), with recaptures of 29 from the Penobscot River and 13 from the Merrimack River. Canadian-origin tags came from the Saint John River, NB (3) and St. Jean, Québec (1). The differences in numbers of recoveries among river systems do not necessarily imply differential exploitation because of the differences in relative numbers of tags at large, and the location and time of sampling. The Working Group noted apparent spatial and temporal clustering of tag recoveries and recommended additional analyses to discern broad-scale migration patterns.

Fourteen Carlin tags from Maine-origin salmon were reported from Canadian fisheries during 1991. Twelve of these tags came from fishermen in SFA 3; tags were also reported from SFAs 4 (1) and 5 (1). Estimates of harvest of USA-origin salmon captured in the 1991 fisheries will not be available until after the 1992 returns to homewaters.

### 6.1.2 Historical data on tag returns and harvest estimates

## Carlin-based estimates for Canada

The Working Group updated the time series of tag returns and harvest estimates of Maine-origin 1SW and 2SW salmon in Newfoundland and Labrador. Tag returns for Maine-origin 1SW salmon can be found in Table 6.1.2.1. There were two tag returns from fisheries prior to 1991, but neither of these tags could be assigned to the fishery in a specific year due to incomplete recovery information.

It was noted that the two tags recovered in SFA 11 in 1990 were reported to be from salmon caught after the season was closed. One tag was from a salmon caught in a cod trap, the other was from a salmon caught with marine angling gear.

Neither the structure of the harvest model nor its parameter values were changed from the previous assessment in Anon. (1991a). Updated values and the new data for the 1991 run size used to calculate the RATIO parameter can be found in Anon. (1992a). Estimates for tags and total run size of 2 SW salmon to Maine rivers, using a fishway efficiency of $85 \%$, are presented in Table 5.1.3.1. For 1991, the estimates of tags and run size were 29 tags and 2,057 fish, respectively. The relatively low number of tags in the run resulted in second lowest RATIO parameter used in the harvest model time series (RATIO for $1991=0.0138$ ), with only 1969 ( 0.0104 ) being lower.

Estimated harvests of 1SW salmon in Newfoundland and Labrador are summarized by year for $85 \%$ fish passage
efficiency in Table 6.1.2.2. The total harvest of 780 Maine-origin salmon in the 1990 fishery was distributed primarily in SFAs 2-4. The 1990 harvest estimate is lower than the 1989 estimate, but is still higher than those observed for the period 1986 through 1988. There was one 2 SW salmon tag return from the 1991 fishery. The tag was recovered in SFA 3 and represented a harvest of 103 fish. Updated values for 1SW:2SW ratios and ratios of harvest to run size of 2 SW salmon are presented in Table 6.1.2.3. The grilse to salmon ratio for the 1989 smolt class is 0.318 which continues the trend of relatively high grilse returns that began in 1986.

Detailed summaries of the tag returns and harvests for 1SW and 2 SW salmon in mainland Canada (Anon., 1990b) are not presented in this assessment since there were no new tag returns for these areas and age groups.

## CWT-based estimates for locations sampled

The Working Group considered harvest estimates of Maine-origin stocks in Newfoundland and Labrador derived from CWT- and Carlin-tagged salmon for areas sampled. Estimates based on Carlin tags for the 1990 fishery were calculated identically to the methods used in other reports of Carlin-derived harvest estimates (Anon., 1987; Anon., 1990a). CWT harvest estimates were computed identically to the methods used in Anon. (1990a). To compute the tag raising factor for the 1990 CWT recoveries the following input data were used:

## 2SW salmon

| Run Estimate | 2057 |
| :--- | ---: |
| Total CWT | 438 |
| CWT at the trap | 334 |
| Angled CWT | 45 |
| Untagged at the Trap | 1343 |
| Untagged Angled | 145 |

From these data, the CWT to run ratio was 0.213 and the raising factor for tags was 4.696 . The estimated numbers of tags for sample strata in the fishery was first raised for non-catch fishing mortality (Anon., 1989b) and then raised to total harvests for the strata. Comparative harvest estimates based on CWT and Carlin tag recoveries were calculated for the communities, statistical sections, and salmon fishing areas sampled. The estimates ranged from 0 fish in some communities to a total of 379 for SFA 4 (Table 6.1.2.4). Coefficients of variation (CV) could not be calculated for most community estimates because scanning exceeded the reported catch in some strata. CVs for salmon fishing area estimates were $6 \%$. The ratio of harvest estimation methods (Carlin/CWT) was greater than 1.0 for the salmon fishing area estimates, but this comparison was made with a recovery of relatively few Carlin tags.

As previously noted in Anon. (1989a), scanning samples of landings that are not sorted by size category and incompatibilities between the scanning samples and catch statistics are potential sources of bias in the harvest model. These potential sources of bias are more problematic for the community level harvest estimates than the pooled estimates for statistical sections or salmon fishing areas.

### 6.1.3 Exploitation rates

No new information on exploitation rates in commercial fisheries was available to the Working Group. Exploitation rates in the recreational fisheries have been estimated using mark/recapture techniques in the Miramichi River (SFA 16), the Margaree River (SFA 18) and the Humber River (SFA 13). Exploitation rates from the Miramichi River and the Margaree River which were reported in Anon. (1991a) had been adjusted for estimates of reporting rate and tag loss. Since reporting rates are difficult to estimate and in some cases, are entirely dependent upon the estimates of angling catch from different sources, it was considered informative to review the unadjusted rates which represent the actual available data for comparison with the adjusted rates.

Atlantic salmon from all three rivers were captured in tidal waters and marked with individually numbered, blue Carlin tags, attached with stainless steel wire. In 1991, a total of 787 small ( $<63 \mathrm{~cm}$ ) and 520 large salmon were marked and released from the three rivers. The unadjusted reporting rates were calculated as the tags returned by anglers to total marks released. Adjusted rates were calculated to account for reporting rate and tag loss. Exploitation rates for large salmon apply to hooked and released fish and may not necessarily apply to retained catches. For the Miramichi River, the exploitation rate based on the estimated angling catch to the estimated returns to the Millbank trap were calculated for comparison. Returns to Millbank trap were estimated using counts of marked and unmarked fish at several counting facilities upstream.

The exploitation rates for early and late runs of small and large salmon in 1991 were similar to those previously observed (Table 6.1.3.1):

- exploitation rate varied on an annual basis in all the rivers,
- exploitation rates on late run small and large salmon are less than on the early run, exploitation rate on large salmon is less than that on small salmon,
- combined reporting rate and tag loss factors tend to be about $50 \%$ in most of the fisheries.

The exploitation rates obtained using the estimates of angling catch to estimated returns to the river for the

Miramichi River are substantially higher than the unadjusted values and are similar to the 0.38 to 0.55 previously estimated. The very high exploitation rate of hooked and released large salmon ( 0.21 to 0.70 ) calculated using the angling catch to total returns estimate probably relates to the difficulty associated with the collection of reliable estimates of hooked and released fish.

### 6.2 Description of Fisheries in United States of America

There were no new management measures instituted in the USA during 1991. The sport catches (number of fish killed), unadjusted for reporting rate, of Atlantic salmon by sea age in Maine rivers, in 1991, were as follows:

| River | umber of Atlantic Salmon |  |  |  | Total$1991$ | Total$1990$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | 2SW | 3SW | PS |  |  |
| St. Croix | 2 | (grilse only) |  |  | 2 | 2 |
| Dennys | 1 | 6 | 0 | 0 | 7 | 33 |
| E. Machias | 2 | 3 | 0 | 0 | 5 | 48 |
| Machias | 2 | 0 | 0 | 0 | 2 | 2 |
| Pleasant | (catch and release only) |  |  |  | 0 | 0 |
| Narraguagus | 1 | 20 | 0 | 1 | 22 | 51 |
| Union | 0 | 0 | 0 | 0 | 0 | 0 |
| Penobscot | 40 | 148 | 1 | 3 | 192 | 431 |
| Ducktrap | 0 | 0 | 0 |  | 0 | 3 |
| Sheepscot | 0 | 4 | 0 | 0 | 4 | 9 |
| Kennebec | 0 | 4 | 0 | 0 | 4 | 46 |
| Saco | 0 | 0 | 0 | 0 | 0 | 16 |
| Other (Marine) | 0 | 0 | 0 | 0 | 0 | 1 |
| Total | 48 | 185 | 1 | 4 | 238 | 642 |
| (1990 Total) | (52) | (562) | (12) | (16) | (642) |  |

Recreational catches of Atlantic salmon were about 63\% lower in 1991 than in 1990. The decreased catch was attributed to smaller runs of salmon and slightly (4\%) lower license sales. The number of salmon caught and released in Maine rivers exceeded the number caught and killed. The catches in USA rivers are believed to be of USA-origin.

The average exploitation rate on salmon on all age classes in the Penobscot River was $11.5 \%$ which is slightly lower then the exploitation rate ( $13.5 \%$ ) observed in 1990. These estimates were based upon the fish passage efficiency ( 0.85 ) and reporting rate ( 0.80 ) adopted by the Working Group in previous years. The exploitation rates on wild salmon were about twice that on hatchery salmon.

|  | Hatchery |  |
| :--- | :--- | :--- |
|  |  | Wild |
| 1SW | 0.16 |  |
| 2SW | 0.08 | 0.30 |
|  |  | 0.17 |

This difference is consistent with previous observations (Anon. 1990a) and related to earlier run timing of wild salmon into the Penobscot River.

### 6.3 Description of Fisheries in France (Islands of St. Pierre and Miquelon)

Catch of salmon for the Islands of St. Pierre and Miquelon in 1991 was 1 t (Table 2.1.1). There were 13 professional fishermen and 37 recreational fishermen in 1989. Tag returns from previous years indicate that salmon of Canadian and USA-origin have been caught in the fisheries of St. Pierre and Miquelon

### 6.4 Evaluation of the Effects of Quota Management Measures Taken in 1990 and 1991 in Newfoundland-Labrador Commercial Fisheries on Stocks and Fisheries in the Commission Area

The Newfoundland and Labrador commercial salmon fishery was under quota management in 1990 and 1991. The quotas, in tonnes, were set by Salmon Fishing Area (SFA) (Table 6.4.1, Figure 3.3.1.1). The fishing season in both 1990 and 1991 began on June 5 and closed in each SFA when the quota was reached or October 15 as in previous years. The quota was controlled by weekly and daily monitoring of landings by Fisheries Officers. In 1990, quotas were attained in 8 SFAs $(4,5,6,8,10$, 11,13 and 14) resulting in closing dates in these fisheries, ranging from June 21 to July 25. In 1991, quotas were attained in 9 SFAs ( $3,5,6,7,8,9,10,11$, and 13) with closing dates ranging from June 29 to September 23. A preliminary evaluation of the effects of the quota management in 1990 was reported in Anon. (1991a,b). That evaluation is up-dated with final catch data and estimates of harvest of 1 SW salmon of Maineorigin in 1990, and preliminary catch data for 1991.

### 6.4.1 Effects on Canadian stocks and fisheries

The quantities of large and small salmon affected by the early closure of the fisheries were evaluated by applying the closure date in each SFA, in 1990 and 1991, to the temporal distribution of the landings in each SFA and year, 1984-89. Thus, the landings in each SFA were divided into two periods. The landings in period 1 represent the total landings which would have occurred if the closure dates in 1990, 1991 had been applied in years 1984-89; whereas, the landings in period 2 represent the amount of landings which would not have been caught. The minimum, maximum and mean percents of the landings in period 2 were calculated and applied to
the landings in the corresponding SFA in 1990 and 1991.

For 1990, the estimated total mean weight (79t) of small salmon is about twice the estimated weight ( 39 t ) of large salmon which were not caught due to the early closure of the fisheries (Table 6.4.1.1). The estimated mean numbers of fish not caught were 41,600 small salmon and 8,600 large salmon (Table 6.4.1.2).

The mean predicted weight of small salmon not caught in 1991 is 21 t and for large salmon is 9 t (Table 6.4.1.1). These weights are equivalent to about 12,600 small salmon and 2,500 large salmon (Table 6.4.1.2).

The minimum and maximum effect of the quotas on the fisheries in all SFAs combined was estimated by: examining the temporal distribution of the total landings, 1984-89, for SFAs closed in 1990 and 1991; selecting the year of lowest (1989) and highest (1985) landings in period 2; and, applying the percent of the landings in period 2 to the total landings in 1990 and 1991. The results are given below.

|  | Not caught all SFAs | Small |  | Large |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | (t) | No. | (t) | No. |
| 1990 | Min | 52 | 27,400 | 27 | 5,700 |
|  | Max | 383 | 174,100 | 97 | 19,700 |
| 1991 | Min | 14 | 7,900 | 4 | 1,000 |
|  | Max | 55 | 32,100 | 24 | 6,000 |

It took longer to reach the quotas in most SFAs in 1991 than in 1990 even though the quotas were similar or lower in 1991 (Table 6.4.1). The total landings were also lower in 1991 suggesting that the total population size in 1991 was less than in 1990. Other factors which may have contributed to the later closing dates were: a delay in migration timing experienced due to the adverse ocean conditions; and, reduced fishing effort at the beginning of the fishing season due to the presence of ice. The estimated average numbers of small $(12,600)$ and large $(2,500)$ salmon not caught in 1991 is about $70 \%$ less than the estimated numbers of small $(41,600)$ and large $(8,600)$ salmon not caught in 1990 (Table 6.4.1.2). In both years the quota had the greatest effect in reducing the numbers of small salmon caught in SFAs 10,11 and 13 (Table 6.4.1.2). The largest reduction in large salmon occurred in SFA 11 in both 1990 and 1991 (Table 6.4.1.2).

In both 1990 and 1991, the quota had a greater effect on proportionally reducing the catch of small salmon than large salmon in most SFAs (Table 6.4.1.1). This difference in reduction was expected because the large salmon tend to migrate earlier along the coast than small
salmon.
Analysis of the description of the composition of the fisheries, 1969-75 by Pippy (1982) indicates that in SFAs 3 to 14 , after July 15, the majority of the salmon in these SFAs were of Newfoundland and Labrador-origin. Some of the salmon not caught due to the closure will be non maturing and would not return to rivers until the following year. Also some of the salmon which were not caught in a SFA due to the closure of the fishery may have been caught in an adjacent fishery which was still open.

There are insufficient data to quantify the composition of the stocks or estimate the additional escapements in the many rivers that may have benefitted from the early closures. Some of the additional salmon entering the rivers due to the early closure would be harvested in the angling fishery. The quantity harvested will vary among river systems due to differing exploitation rates (Anon 1990b) and annual variation in stock size. There are presently no measures of total stock size.

### 6.4.2 Effects on USA stocks

Sufficient information was only readily available to assess the effects of the 1990 quotas on the harvest of the 1SW salmon of Maine-origin. Impacts of the 1991 quotas on 1SW Maine-origin salmon can not be assessed until 1992 when estimates of the harvest of Maine-origin salmon in the 1991 fisheries will be available. The numbers of 1SW Maine-origin salmon which were not caught due to the earlier closures were estimated by examining the estimates of the weekly harvests, June 5 to October 15, each year, 1984-89; the harvests which had not been assigned to a SFA or week were not included. The harvest in each SFA each year, 19841989, (Table 6.4.2.1) was divided into 2 periods using the date the fisheries were closed in each SFA in 1990. The percent of the total harvest (period $1+2$ ) which was caught in period 2 was computed by SFA and SFAs combined. The mean percent is the total number of salmon caught in period 2 divided by the total catch in period $1+2$. The mean percentage of the harvest, in period 2, for all SFAs combined, over years, was applied to the total estimates of harvest of 1SW salmon of Maine-origin, 1990, in the 8 SFAs which were closed prior to October 15 to give an estimate of the numbers of salmon not killed due to early closure of the fisheries in 1990. The above calculations were repeated using the date the fisheries were closed in each SFA in 1991 to compute the average number of salmon and the percentage of the total harvest which would not have been caught, 1984-89.

The percent of the harvest of 1SW Maine-origin salmon which would not have been harvested in the Newfound-land-Labrador commercial fisheries if the closure date
for each SFA in 1990 was applied in 1984-89 varied from $0 \%$ to $100 \%$ depending on SFA and year (Table 6.4.2.2). The small numbers of salmon harvested and the high variability in the percent of the harvest which could be affected by the closure makes it difficult to estimate the effects of the closure on Maine-origin salmon. The average percentage of the harvest which would not have been caught in SFAs 4, 5, 6, 8, 10, 11, 13 , and 14 for all years $1984-89$ is $42 \%$ (S.D. 22.81) (Table 6.4.2.1). An estimate of 221 (95\% C.L. 65 589) 1SW Maine-origin salmon would not have been caught due to the closure of the fisheries in these SFAs. It is not possible to determine how many additional 2SW salmon actually returned to Maine rivers in 1991 due to the annual variability in returns to the rivers (Anon, 1991b).

The mean percent harvest on 1SW Maine-origin salmon which would not have been caught if the 1991 closure dates were in effect, 1984-89, is $16 \%$ (Table 6.4.2.2), $63 \%$ less than using the 1990 closure dates (Table 6.4.2.1). This difference suggests that the quotas in 1991 were less effective in proportionally reducing the harvest than in 1990.

### 6.5 By-catches of Fish, Birds and Marine Mammals in Salmon Drift-net Fisheries

The Working Group is not aware of any legal or illegal drift-net fisheries for salmon in the North American Commission area.

## 7 INVENTORY OF PARASITES AND DISEASES BY COUNTRY

The Working Group reviewed the Report of the Working Group on Pathology and Diseases of Marine Organisms (WGPDMO, Anon. 1992e). The WGPDMO was asked by ICES to prepare an inventory of parasites and diseases of wild and reared salmon by country and to report to the North Atlantic Salmon Working Group.

WGPDMO presented qualitative data from Faroes, France, Iceland, Ireland, Norway and Scotland (Table 7.1). A large number of different diseases was identified originating from virus, bacteria, fungi, protozoans, monogeans, trematodes, cestodes, nematodes, achanthocephalans, crustaceans, molluscans and leeches. It was noted that there were considerable problems with compiling and presenting these data. The WGPDMO expressed that the inventory was incomplete and potentially misleading, and stressed that these data should only be used with full awareness of the following constraints:
(1) The information more likely reflects the amount of research activities in this field rather than the actual situation.
(2) Many reports are single observations often in individual fish.
(3) The record of a disease/parasite does not necessarily mean that it is still present.
(4) Unless a disease/parasite has been specifically looked for, its absence from the list for a particular country, or in farmed or wild fish, cannot be taken as evidence of its absence.
(5) The compilation of the data were not made on the same basis in all countries.
(6) There is controversy regarding the specific identification of some pathogens/parasites and, consequently, records of a particular example from different countries may in reality refer to different infections.

The Working Group noted the limitations of the inventory presented and endorsed the recommendation given by the WGPDMO: "NASCO should take note of the limitations of this inventory and its vulnerability to misinterpretation and the WGPDMO urges caution in its use. For example, the circumstances of pathogen/parasite detection and diagnosis must be taken into account in assessing the relative degrees of significance of the listed examples."

## 8 REPORT OF THE WORKSHOP ON SALMON ASSESSMENT METHODOLOGY

The Workshop on Salmon Assessment Methodology (Anon., 1992c) met in Dublin (Ireland) from the 2-4 March 1992 with the main task of reviewing and reporting on assessment methodologies. A further task was to standardize the ageing nomenclature used for Baltic and Atlantic salmon.

Models were used to assess stocks in 3 main areas, the Baltic and the Northeast and Northwest Atlantic. The questions to be addressed and the data available in these areas dictates the types of models that can be used. The models currently in use or under consideration were described by assessment area.

A full description and specification of each model was provided and sensitivity analysis was provided where appropriate. A summary chapter provides a quick reference guide for the models and details some of the perceived advantages and disadvantages associated with them.

To help in understanding the underlying need for the various models, an overview of the life history of the salmon, the fisheries, and the general management philosophy by area was given. Some terminology was
also defined where the possibility for confusion existed.
With regard to the question of standardizing the ageing nomenclature used for Baltic and Atlantic salmon, the participants found that there were valid reasons for using the existing nomenclature mainly associated with the differences in the times of the execution of the fisheries. It was felt that a change would lead to considerable confusion and could not be recommended. A table showing equivalent ages was provided to help in comparing ages in each area.

## 9 PRODUCTION OF FARM SALMON

The production of farmed salmon by several countries, as reported to the Working Group (Table 9.1), was $217,569 \mathrm{t}$ in 1991. This represents a small decrease of $6,661 \mathrm{t}$ on the 1990 figure (Figure 9.1), mainly due to lower production in Norway. Total farm production was more than 50 times the nominal wild catch.

## 10 COMPILATION OF TAG RELEASES AND FIN-CLIP DATA FOR 1991

Data were provided by Working Group members on the prescribed form and have been compiled as a separate report (Anon., 1992d). In excess of 1.76 million microtags (CWTs) and 0.33 million external tags were applied to Atlantic salmon released in 1991 (Table 10.1). In addition, 1.63 million salmon were finclipped, 1.45 million with adipose finclip only. Thus, more than 3.72 million marked fish were released.

## 11 RESEARCH

### 11.1 Progress on Data Requirements and Research Needs

The Working Group reviewed the requirements for future meetings (Section 13.3) and progress on data requirements and research needs (Section 13.1) in Anon. (1991a). The progress made to date is noted for all items of Section 13.3 (Anon. 1991a) and those outstanding items (no. 8-11 below) from Section 13.1 (Anon. 1991a) which remain relevant.

1. The Working Group expressed concern over the fact that a substantial portion of the CWT estimate of Maine-origin fish harvested in the West Greenland fishery was based on a limited number of tag recoveries and recommended that the method of stratification of CWTs over weeks and areas in which there was no scanning be re-investigated.

## Task completed.

2. The Working Group endorsed the recommendations of the Workshop on Identification of Fish Farm Escapees and Wild Salmon and also noted that further research on otoliths may lead to the development of techniques to identify fish escaping at the smolt stage. The Working Group also suggested that methods of artificially marking otoliths of farmed fish, such as with oxytetracycline or by the deliberate introduction of growth zones (e.g., by temperature variation) should be investigated.

## No progress.

3. The analysis of the reporting rate for Maineorigin Carlin tags in relation to catch-per-unit-effort for the first two weeks of the West Greenland fishery should be made for other tagging programmes.

## No progress.

4. Error estimates for the Carlin harvest estimates of Maine origin salmon in Greenland should be derived if possible.

## No progress.

5. Studies should be conducted to investigate the incidence of fish farm escapees in the catch at West Greenland.

## Task completed.

6. With respect to the 'bottom-up' run-reconstruction models for countries of the North-East Atlantic, further attempts should be made to refine each national model before 1992. Further, it was agreed that all members should review their tagging experiments and consider what additional data would be required to improve the reliability of their national model.

## Little progress reported at Study Group.

7. The bias-corrected catch-weighting technique by size category is the preferred method for determining the number of North American and European salmon captured in West Greenland. Existing catch-weighted estimates (by size category) are not bias-corrected and the Working Group recommended that this be done for as many years as possible for the next meeting.

## Some progress made.

8. The Working Group requires further work on biochemical techniques, incorporating additional samples, and suggests that this methodology merits further investigations for country of origin.

No new progress reported.
9. The Working Group noted that no effort data were available from the West Greenland fishery in 1989 and recommends that in future years estimates of effort be provided for the 'small' and 'large' boat components, and for individual fishermen who might cooperate in the daily completion of catch/effort logbooks.

No new information tabled, but there is an expectation that data will be tabled in the future.
10. The possibility of obtaining data on weight categories recorded by the factories at West Greenland should be investigated.

## No new information obtained.

11. The Working Group recommends that for 1990-1991, countries should develop run reconstruction models of their stocks for discussion at Study Groups and input to a North Atlantic model at the Working Group.

Significant progress was again made in run reconstruction modelling (Section 5) and the Working Group recommends continued development.

### 11.2 Progress on Recommendations from 1991

1. The Working Group considered a series of indicators of stock status and sustainability which could be used to advise managers of Atlantic salmon stocks. Members reported on numerous time series of data on catches, spawning escapement, stock-specific exploitation rates, etc., which have not been previously utilized by the Working Group in the preparation of their report. Trends over time in raw, normalized, smoothed-normalized (3-year moving averages) and river-combined smoothed-normalized data provided graphic outputs which captured trends potentially useful in future Working Group reports and reports of ACFM. The Working Group recommended that participants bring these additional data bases in electronic format for review at future meetings. The Working Group chairman, in collaboration with the chairmen of the two Study Groups, will undertake to standardize the format prior to the 1992 meetings.

## Considerable progress made.

2. The Chairmen of the North Atlantic Salmon Working Group and the Baltic Salmon and Trout Working Group discussed a number of topics of interest to both Groups with the intention of holding a Workshop.

Workshop on salmon assessment methodology held March 2-4, 1992.
3. Tag recoveries are crucial for the improvement of models being developed to assess the West Greenland and Faroes fisheries. The Working Group recommends that an effort be made to improve and increase screening for tags in the catches, especially in West Greenland.

## Progress reported.

4. The Working Group was encouraged by the substantial advances in the discrimination of continent of origin of salmon at Greenland. It was recommended that the investigations continue and that the technique with the lowest miss-classification and error rate be used to discriminate continent of origin of the salmon in the 1991 fishery at Greenland. Specific recommendations are:
i) Ongoing investigations into scale character variables require a larger data set; the inter-circuli spacings data set looks the most promising of the 5 data sets examined but the analysis should be restricted to single combinations of variables so as to avoid collinearity; pattern recognition and the efficacy of Fourier coefficients should be investigated.
ii) Work should continue on discriminant and maximum likelihood techniques, especially the possible use of a quadratic discriminant analysis which may perform better than the linear discriminant analysis.
iii) Samples should be collected to establish the reliability of the rDNA technique to identify salmon collected at Greenland as North American or European.
iv) Investigation of neural networks should continue.

## Progress reported on neural networks.

5. Further to the development of techniques to determine the state of maturity of 1 SW salmon caught in fisheries it is recommended that:
i) samples be obtained to test the hypothesis that gonadal and somatic growth are equal;
ii) data from the Miramichi River and Norwegian cage experiments be used to test the assumption that there are no differences in GSI between stocks or over time and that Conne estuary and Labrador Sea samples are representative of maturing and non-maturing salmon, respectively;
iii) known maturity samples be collected over time and maturity rates be catch-weighted to test
the assumption that there are no timing differences in GSI and maturity over the weeks of the fishery.

## No progress made.

6. The Working Group was encouraged by new progress into a "top-down" run reconstruction model which was used to constrain the bounds on possible exploitation rates in the West Greenland fishery. The Working Group recommended that work should be done to analytically derive the range of exploitation rates implied by the set of constraints and to examine longer time series of data,particularly for periods prior to major fishery restrictions in Greenland and Canada. Such analyses might detect the effects of management measures, trends in recruitment and patterns of exploitation.

## Significant progress reported at meeting.

7. In addition to the experimental design proposed by the Study Group on Genetic Risks, the Working Group suggests that several less comprehensive experiments might be carried out. These would examine relative performance of farmed and wild salmon at various life stages and would contribute specific information. There is scope for laboratory-based inter-comparisons of genetic and behaviourial interactions between farmed and wild salmon. These would be viewed as complementary to field experiments and could yield useful information that could not otherwise be obtained.

## No progress reported.

### 11.3 Requirements for Future Meetings

1. Nominal catches by country (Table 2.1.1) do not in themselves provide insight into their potential to contain intercepted salmon of another country. It was agreed that nominal catch of salmon reported by country would be more meaningful if the existing time series and future reporting could be separated into coastal and estuarine/freshwater components.
2. The Working Group developed and applied a continental run reconstruction model to estimate abundance of North American salmon at West Greenland. Abundance estimates for North American stocks were then used to define a range of estimates of pre-fishery abundance and exploitation rates. The interactive effects between abundance and exploitation rates at West Greenland were then modelled in relation to achievement of North American spawning targets. The Working Group recommended that: i) with the potential of the approach to interpret historical patterns of abundance and exploitation rates, that total return and catch data be assembled
for fishery years prior to 1983, ii) that further analyses be carried out to examine the different contributions of stocks from different areas to the West Greenland fishery, iii) that future refinement of this approach should include assessment of the relative risks associated with errors in the prediction of pre-fishery abundance and application of the model to groups of stocks (e.g., river age 4 and older, and river age 3 and younger).
3. The Working Group also examined information from Canada and West Greenland which may provide an estimate of pre-fishery and in-season abundance at West Greenland. The Working Group recommended that: i) 1SW data from fisheries, river counts and traps in Europe be examined for pre-season predictors, and, ii) that all available daily catch data (by small and large boats) from the West Greenland fishery be examined with more rigorous statistical procedures e.g., probability distribution functions for an in-season predictor.

### 11.4 Research Programs

1. The identification of North American and European fish in West Greenland in 1991 was again hindered by the lack of a suitable test sample of fish of known origin and the high misclassification rates associated with the historical database used to form the discriminant function. As in earlier years, scale samples should be collected from 2SW salmon in home-waters in 1992 and forwarded to D.Reddin, Canada.
2. The Working Group noted the importance of continuing the scientific research and monitoring in the Faroes fishery area and expressed an interest in seeing the results from the current research program.

### 11.5 Future Meeting

The Working Group enthusiastically congratulated John Browne, his colleagues and staff from the Fisheries Research Centre and many Irish friends on their enthusiastic provision of cheerfulness, cooperation and a splendid meeting facility and working environment. Go raibh maith agibh!

The Working Group recommended, that unless invitations were tendered to, and acceptable to the General Secretary and the Chairman of Working Group, that the Working Group meet at ICES headquarters, in 1993, preferably in late March.

## 12 RECOMMENDATIONS

1. The Working Group endorsed the recommendations of the Study Group on the North American Salmon Fisheries, the Study Group on the Norwegian Sea and Faroes Salmon Fishery, and the Workshop on Salmon Assessment Methodology (Appendix 4).

If there is a requirement by ICES for the Study Group on North American Salmon Fisheries to meet in 1993, the Working Group encouraged national agencies to: consider the adoption of graphic methods to depict measures of central tendency, trends etc.; investigate stock-recruit relationships for naturally spawning fish in the Penobscot and Saint John rivers and to examine forecast models for MSW salmon in an attempt to explain observed recent decreases in the numbers of MSW salmon and increases in the numbers of 1 SW salmon returning to some Canadian rivers.

The Study Group on the Norwegian Sea and Faroes Salmon Fishery should be renamed the Study Group on North East Atlantic Salmon Fisheries, thereby recognizing the discontinuation of commercial salmon fishing by Faroese boat owners within the Faroes EEZ through the 1993/94 season and the continuing need to collate information on homewater fisheries. The new Study Group should meet at ICES headquarters for 4 days immediately prior to the meeting of the Working Group on North Atlantic Salmon to compile relevant data including that from 1989 and 1990 smolt releases for the run-reconstruction model.
2. The Working Group concluded that the neural network analysis may offer an alternative to the discriminant analysis of separating stocks of North American and European salmon at West Greenland. The Working Group recommends that the neural network methodology undergo further testing using simulation data sets run parallel with discriminant function analysis. Specifically, the neural network should be evaluated with respect to its: i) sensitivity of classification success with respect to the order of input observations, ii) sensitivity of classification success to category boundaries applied to the input parameters, iii) ability to discriminate simulated samples drawn from underlying distributions with known parameters, iv) relationship to traditional statistical methods, particularly log linear models for categorical data.
3. A considerable number of USA tags have recently been returned from West Greenland without year-of-capture data. Virtually all tags could be safely assigned a year of capture on the basis of smolt-release information. The Working Group suggested an investigation of the assumptions used to set non-reporting rates that would be violated if these tags were now used in the Carlin tag method of estimating Maine origin salmon at West Greenland.
4. The Working Group noted that increasingly larger proportions of the landings at West Greenland are taken by boats smaller than 30 feet. The Working Group recommended that changes in fleet characteristics be considered as a means of adjusting non-catch fishing mortality for use in the proportional harvest model.
5. The Working Group endorsed a recommendation for a Working Group memeber, L. W. Stolte (not in attendance), to develop and implement a video/slide exchange program between interested parties in Atlantic salmon producing countries. The exchange could focus on techniques and installations for salmon culture, fish passage, harvest, juvenile stocking, etc. The content could be reviewed in the context of a scientific critique. Working Group participants volunteering to be the contact person(s) for their country are astericked in Appendix 5.
6. The Working Group noted recent declines in marine survival of Atlantic salmon. No data were formally tabled at the Working Group that related survival and causality, although it was apparent that several investigations were on-going. The Working Group encouraged these investigations, especially those that examined causality of survival patterns of stock complexes at large in the North Atlantic. Investigations should include the use of historical data to determine trends in stock status, comparability of trends between
stocks of different river systems that can be ascribed similar migration patterns, and the examination of biological/environmental factors that could explain variation in abundance.
7. The Working Group made considerable progress in efforts to analyze and present time series of data so as to identify trends descriptive of the status of stocks. Additional data sets were made available which had not previously been utilized. Much data remains to be examined. In addition to techniques used to examine/depict trend analysis in this report, the Working Group recommended consideration of the Lowess technique and the diagnostics of "route regression" to assist in the identification, grouping and testing of similarities/differences of trends in time-series data.
8. The Working Group examined the adequacy of the sampling program at West Greenland and recommended that sampling be maintained within the 3 NAFO Divisions and that the duration of sampling within one or two of these locations be extended by one or two weeks.

Table 2.1.1 Nominal catch of SALMON by country (in tonnes round fresh weight), 1960-1991 (1991 provisional figures)

| Year | Canada <br> (4) | Den. | Faroes | Finland | France | East <br> Grld. | West Grld. | Iceland | Ireland (1) | Norway (3) | Russia | $\begin{aligned} & \text { St. P } \\ & \text { \& M. } \end{aligned}$ | Sweden <br> (W. C.) | UK E. \& W. | UK Scotland | UK $\text { N.I. }(1,2)$ | USA | Others (5) | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | 1636 | - | - | - | - | - | 60 | 100 | 743 | 1659 | 1100 | - | 40 | 283 | 1443 | 139 | 1 | - | 7204 |
| 1961 | 1583 | - | - | - | - | - | 127 | 127 | 707 | 1533 | 790 | - | 27 | 232 | 1185 | 132 | 1 | - | 6444 |
| 1962 | 1719 | - | - | - | - | - | 244 | 125 | 1459 | 1935 | 710 | - | 45 | 318 | 1738 | 356 | 1 | - | 8650 |
| 1963 | 1861 | - | - | - | - | - | 466 | 145 | 1458 | 1786 | 480 | - | 23 | 325 | 1725 | 306 | 1 | - | 8576 |
| 1964 | 2069 | - | - | - | - | - | 1539 | 135 | 1617 | 2147 | 590 | - | 36 | 307 | 1907 | 377 | 1 | - | 10725 |
| 1965 | 2116 | - | - | - | - | - | 861 | 133 | 1457 | 2000 | 590 | - | 40 | 320 | 1593 | 281 | 1 | - | 9392 |
| 1966 | 2369 | - | - | - | - | - | 1370 | 106 | 1238 | 1791 | 570 | - | 36 | 387 | 1595 | 287 | 1 | - | 9750 |
| 1967 | 2863 | - | - | - | - | - | 1601 | 146 | 1463 | 1980 | 883 | - | 25 | 420 | 2117 | 449 | 1 | - | 11948 |
| 1968 | 2111 | - | 5 | - | - | - | 1127 | 162 | 1413 | 1514 | 827 | - | 20 | 282 | 1578 | 312 | 1 | 403 | 9755 |
| 1969 | 2202 | - | 7 | - | - | - | 2210 | 133 | 1730 | 1383 | 360 | - | 22 | 377 | 1955 | 267 | 1 | 893 | 11540 |
| 1970 | 2323 | - | 12 | - | - | - | 2146 | 195 | 1787 | 1171 | 448 | - | 20 | 527 | 1392 | 297 | 1 | 922 | 11241 |
| 1971 | 1992 | - | - | - | - | - | 2689 | 204 | 1639 | 1207 | 417 | - | 18 | 426 | 1421 | 234 | 1 | 471 | 10719 |
| 1972 | 1759 | - | 9 | 32 | 34 | - | 2113 | 250 | 1804 | 1568 | 462 | - | 18 | 442 | 1727 | 210 | 1 | 486 | 10915 |
| 1973 | 2434 | - | 28 | 50 | 12 | - | 2341 | 256 | 1930 | 1726 | 772 | - | 23 | 450 | 2006 | 182 | 2.7 | 533 | 12746 |
| 1974 | 2539 | - | 20 | 76 | 13 | - | 1917 | 225 | 2128 | 1633 | 709 | - | 32 | 383 | 1708 | 184 | 0.9 | 373 | 11941 |
| 1975 | 2485 | - | 28 | 76 | 25 | - | 2030 | 266 | 2216 | 1537 | 811 | - | 26 | 447 | 1621 | 164 | 1.7 | 475 | 12209 |
| 1976 | 2506 | - | 40 | 66 | 9 | $<1$ | 1175 | 225 | 1561 | 1530 | 772 | 2.5 | 20 | 208 | 1019 | 113 | 0.8 | 289 | 9536 |
| 1977 | 2545 | - | 40 | 59 | 19 | 6 | 1420 | 230 | 1372 | 1488 | 497 | - | 10 | 345 | 1160 | 110 | 2.4 | 192 | 9495 |
| 1978 | 1545 | - | 37 | 37 | 20 | 8 | 984 | 291 | 1230 | 1050 | 476 | - | 10 | 349 | 1323 | 148 | 4.1 | 138 | 7650 |
| 1979 | 1287 | - | 119 | 26 | 10 | $<1$ | 1395 | 225 | 1097 | 1831 | 455 | - | 12 | 261 | 1076 | 99 | 2.5 | 193 | 8089 |
| 1980 | 2680 | - | 536 | 34 | 30 | $<1$ | 1194 | 249 | 947 | 1830 | 664 | - | 17 | 360 | 1134 | 122 | 5.5 | 277 | 10080 |
| 1981 | 2437 | - | 1025 | 44 | 20 | <1 | 1264 | 163 | 685 | 1656 | 463 | - | 26 | 493 | 1233 | 101 | 6 | 313 | 9929 |
| 1982 | 1798 | - | 865 | 54 | 20 | <1 | 1077 | 147 | 993 | 1348 | 354 | - | 25 | 286 | 1092 | 132 | 6.4 | 437 | 8634 |
| 1983 | 1424 | - | 678 | 57 | 16 | $<1$ | 310 | 198 | 1656 | 1550 | 507 | 3 | 28 | 429 | 1221 | 187 | 1.3 | 466 | 8731 |
| 1984 | 1112 | - | 628 | 44 | 25 | <1 | 297 | 159 | 829 | 1623 | 593 | 3 | 40 | 345 | 1013 | 78 | 2.2 | 101 | 6892 |
| 1985 | 1133 | - | 566 | 49 | 22 | 7 | 864 | 217 | 1595 | 1561 | 659 | 3 | 45 | 361 | 913 | 98 | 2.1 | - | 8095 |
| 1986 | 1559 | - | 530 | 38 | 28 | 19 | 960 | 310 | 1730 | 1598 | 608 | 2.5 | 54 | 430 | 1271 | 109 | 1.9 | - | 9248 |
| 1987 | 1784 | - | 576 | 49 | 27 | <1 | 966 | 222 | 1239 | 1385 | 564 | 2 | 47 | 302 | 922 | 56 | 1.2 | - | 8142 |
| 1988 | 1311 | - | 243 | 34 | 32 | 4 | 893 | 396 | 1874 | 1076 | 419 | 2 | 40 | 395 | 882 | 114 | 0.9 | - | 7716 |
| 1989 | 1139 | - | 364 | 52 | 14 | $<1$ | 337 | 278 | 1079 | 905 | 359 | 2 | 29 | 296 | 895 | 142 | 1.7 | - | 5893 |
| 1990 | 911 | 13 | 315 | 59 | 15 | $<1$ | 227 | 426 | 586 | 930 | 315 | 2 | 33 | 338 | 624 | 94 | 2.4 | - | 4890 |
| 1991 | 679 | 3.3 | 95 | 69 | 13 | 4 | 438 | 519 | 422 | 885 | 215 | 1 | 38 | 199 | 395 | 55 | 0.8 | - | 4031 |

1. Catch on River foyle allocated $50 \%$ Ireland and $50 \%$ Northern Ireland
2. Not including angling catch (mainly grise).
3. Before 1966 , sea trout and sea charr included ( $5 \%$ total).
4. Includes estimates of some local sales and by-catch
5. Includes catches in Norwegian Sea by vessels from Denmark, Sweden, Germany, Norway and Finland

Table 2.1.2 Nominal catch of SALMON in homewaters by country (in tonnes round fresh weight), 1960-1991 (1991 provisional figures). $S=$ Salmon (2SW or MSW fish). $G=$ Grilse (1SW fish). $T=S+G$.

|  | Canada(5) |  |  | Finland |  |  | $\begin{array}{\|l\|} \hline \text { France } \\ \hline \mathrm{T} \\ \hline \end{array}$ | $\begin{array}{\|l} \text { lice- } \\ \text { land } \end{array}$ | Ireland(2) |  |  | Norway(4) |  |  | Russia$T$ | Swe- <br> den <br> (w.C.) <br> $T$ | $\begin{array}{\|l} \text { UK } \\ \text { E.\&W. } \\ \hline \mathrm{T} \\ \hline \end{array}$ | UK <br> N.I. <br> $(2,3)$ <br> $T$ | UKScotland |  |  | $\begin{array}{\|l\|} \hline \text { USA } \\ \hline \mathrm{T} \\ \hline \end{array}$ | Total <br> (6) <br> $T$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Lg | Sm | T | S | G | T |  |  | S | G | T | S | G | T |  |  |  |  | Lg | Sm | T |  |  |
| 1960 |  |  | 1636 |  |  | - | - | 100 | - | - | 743 | - | - | 1659 | 1100 | 40 | 283 | 139 | 971 | 472 | 1443 | 1 | 7144 |
| 1961 |  |  | 1583 | - |  |  | - | 127 | - | - | 707 | - | - | 1533 | 790 | 27 | 232 | 132 | 811 | 374 | 1185 | 1 | 6317 |
| 1962 |  |  | 1719 |  |  |  | - | 125 | - |  | 1459 | - | - | 1935 | 710 | 45 | 318 | 356 | 1014 | 724 | 1738 | 1 | 8406 |
| 1963 |  |  | 1861 |  |  |  |  | 145 | - | - | 1458 | - | - | 1786 | 480 | 23 | 325 | 306 | 1308 | 417 | 1725 | 1 | 8110 |
| 1964 |  |  | 2069 |  |  |  | - | 135 | - |  | 1617 | - | - | 2147 | 590 | 36 | 307 | 377 | 1210 | 697 | 1907 | 1 | 9186 |
| 1965 |  |  | 2116 |  |  |  |  | 133 | - |  | 1457 | - | - | 2000 | 590 | 40 | 320 | 281 | 1043 | 550 | 1593 | 1 | 8531 |
| 1966 |  |  | 2369 |  |  |  | - | 106 | - |  | 1238 | - | - | 1791 | 570 | 36 | 387 | 287 | 1049 | 546 | 1595 | 1 | 8380 |
| 1967 |  |  | 2863 | - |  |  | - | 146 | - | - | 1463 | - | - | 1980 | 883 | 25 | 420 | 449 | 1233 | 884 | 2117 | 1 | 10347 |
| 1968 | - |  | 2111 | - | - | - | - | 162 | - | - | 1413 | - | - | 1514 | 827 | 20 | 282 | 312 | 1021 | 557 | 1578 | 1 | 8220 |
| 1969 | - |  | 2202 | , |  | - | - | 133 | - | - | 1730 | 801 | 582 | 1383 | 360 | 22 | 377 | 267 | 997 | 958 | 1955 | 1 | 8430 |
| 1970 | 1562 | 761 | 2323 | - | - | - | - | 195 | - | - | 1787 | 815 | 356 | 1171 | 448 | 20 | 527 | 297 | 775 | 617 | 1392 | 1 | 8161 |
| 1971 | 1482 | 510 | 1992 |  |  | - | - | 204 | - | - | 1639 | 771 | 436 | 1207 | 417 | 18 | 426 | 234 | 719 | 702 | 1421 | 1 | 7559 |
| 1972 | 1201 | 558 | 1759 |  | - | 32 | 34 | 250 | 200 | 1604 | 1804 | 1064 | 514 | 1578 | 462 | 18 | 442 | 210 | 1013 | 714 | 1727 | 1 | 8317 |
| 1973 | 1651 | 783 | 2434 |  |  | 50 | 12 | 256 | 244 | 1686 | 1930 | 1220 | 506 | 1726 | 772 | 23 | 450 | 182 | 1158 | 848 | 2006 | 2.7 | 9844 |
| 1974 | 1589 | 950 | 2539 |  |  | 76 | 13 | 225 | 170 | 1958 | 2128 | 1149 | 484 | 1633 | 709 | 32 | 383 | 184 | 912 | 716 | 1628 | 0.9 | 9551 |
| 1975 | 1573 | 912 | 2485 |  |  | 76 | 25 | 266 | 274 | 1942 | 2216 | 1038 | 499 | 1537 | 811 | 26 | 447 | 164 | 1007 | 614 | 1621 | 1.7 | 9676 |
| 1976 | 1721 | 785 | 2506 |  |  | 66 | 9 | 225 | 109 | 1452 | 1561 | 1063 | 467 | 1530 | 772 | 20 | 208 | 113 | 522 | 497 | 1019 | 0.8 | 8030 |
| 1977 | 1883 | 662 | 2545 | - | - | 59 | 19 | 230 | 145 | 1227 | 1372 | 1018 | 470 | 1488 | 497 | 10 | 345 | 110 | 639 | 521 | 1160 | 2.4 | 7837 |
| 1978 | 1225 | 320 | 1545 | - | - | 37 | 20 | 291 | 147 | 1082 | 1229 | 668 | 382 | 1050 | 476 | 10 | 349 | 148 | 781 | 542 | 1323 | 4.1 | 6482 |
| 1979 | 705 | 582 | 1287 |  | - | 26 | 10 | 225 | 105 | 922 | 1027 | 1150 | 681 | 1831 | 455 | 12 | 261 | 99 | 598 | 478 | 1076 | 2.5 | 6312 |
| 1980 | 1763 | 917 | 2680 | - | - | 34 | 30 | 249 | 202 | 745 | 947 | 1352 | 478 | 1830 | 664 | 17 | 360 | 122 | 851 | 283 | 1134 | 5.5 | 8073 |
| 1981 | 1619 | 818 | 2437 | - | - | 44 | 20 | 163 | 164 | 521 | 685 | 1189 | 467 | 1656 | 463 | 26 | 493 | 101 | 834 | 389 | 1223 | 6 | 7317 |
| 1982 | 1082 | 716 | 1798 | - | - | 54 | 20 | 147 | 63 | 930 | 993 | 985 | 363 | 1348 | 364 | 25 | 286 | 132 | 596 | 496 | 1092 | 6.4 | 6265 |
| 1983 | 911 | 513 | 1424 | - | - | 57 | 16 | 198 | 150 | 1506 | 1656 | 957 | 593 | 1550 | 507 | 28 | 429 | 187 | 672 | 549 | 1221 | 1.3 | 7274 |
| 1984 | 645 | 467 | 1112 | - | - | 44 | 25 | 159 | 101 | 728 | 829 | 995 | 628 | 1623 | 593 | 40 | 345 | 78 | 504 | 509 | 1013 | 2.2 | 5863 |
| 1985 | 540 | 593 | 1133 | - | - | 49 | 22 | 217 | 100 | 1495 | 1595 | 923 | 638 | 1561 | 659 | 45 | 361 | 98 | 514 | 399 | 913 | 2.1 | 6655 |
| 1986 | 779 | 780 | 1559 | 28 | 10 | 38 | 28 | 310 | 136 | 1594 | 1730 | 1042 | 556 | 1598 | 608 | 54 | 430 | 109 | 745 | 526 | 1271 | 1.9 | 7737 |
| 1987 | 951 | 833 | 1784 | 35 | 14 | 49 | 27 | 222 | 127 | 1112 | 1239 | 894 | 491 | 1385 | 564 | 47 | 302 | 56 | 503 | 419 | 922 | 1.2 | 6598 |
| 1988 | 633 | 677 | 1310 | 26 | 8 | 34 | 32 | 396 | 141 | 1733 | 1874 | 656 | 420 | 1076 | 419 | 40 | 395 | 114 | 501 | 381 | 882 | 0.9 | 6573 |
| 1989 | 590 | 549 | 1139 | 17 | 35 | 52 | 14 | 278 | 132 | 947 | 1079 | 469 | 436 | 905 | 359 | 29 | 296 | 142 | 464 | 431 | 895 | 1.7 | 5190 |
| 1990 | 486 | 425 | 911 | 24 | 35 | 59 | 15 | 426 | - | - | 586 | 545 | 385 | 930 | 315 | 33 | 338 | 94 | 423 | 201 | 624 | 2.4 | 4333 |
| 1991(1) | 361 | 318 | 679 | - | - | 69 | 13 | 519 | - | - | 422 | 538 | 347 | 885 | 215 | 38 | 199 | 55 | 242 | 154 | 396 | 0.8 | 3491 |

$\mathrm{S}=$ salmon (2SW or MSW fish), $\mathrm{G}=\mathrm{grilse}$ (1SW fish). $\mathrm{T}=\mathrm{S}+\mathrm{G} . \mathrm{Sm}=$ small. Lg=large.

1. Provisional figures
2. Catch on River Foyle allocated $50 \%$ lreland and $50 \% \mathrm{~N}$. Ireland.
3. Not including angling catch (mainly griise).
4. Before 1966, sea trout and sea charr included ( $5 \%$ total).
5. Includes estimates of some local sales and by-catch, some fish in "G" column are non-maturing.

Table 2.2.1 Reported catch of SALMON in numbers and weight in tonnes (round fresh weight). Catches reported for 1991 are provisional. Some countries divide ISW from MSW salmon based on weight.

| Country | Year | 1SW |  | 2SW |  | 3SW |  | 4SW |  | 5SW |  | MSW! |  | PS |  | Tota 1 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. | Wt | No. | Wt | No. | Wt | No. |  | No. | Wt | No. | Wt | No. |  | No. | Wt |
| Canada | 1982 | 358,000 | 716 | - | - | - | - | - | - | - | - | 240,000 | 1,082 | - | - | 598,000 1 | 1,798 |
|  | 1983 | 265,000 | 513 | - | - | - | - | - | - | - | - | 201,000 | 911 | - | - | 466,000 1 | 1,424 |
|  | 1984 | 234,000 | 467 | - | - | - | - | - | - | - | - | 143,000 | 645 | - | - | 377,000 1 | 1,112 |
|  | 1985 | 333,084 | 593 | - | - | - | - | - | - | - | - | 122,621 | 540 | - | - | 455,705 | 1,133 |
|  | 1986 | 417,269 | 780 | - | - | - | - | - | - | - | - | 162,305 | 779 | - | - | 579,5741 | 1,731 |
|  | 1987 | 435,799 | 833 | - | - | - | - | - | - | - | - | 203,731 | 951 | - | - | 639,530 1 | 1,784 |
|  | 1988 | 372,178 | 677 | - | - | - | - | - | - | - | - | 137,637 | 633 | - | - | 509,815 | 1,311 |
|  | 1989 | 304, 620 | 549 | - | - | - | - | - | - | - | - | 135,484 | 590 | - | - | 440,104 1 | 1,139 |
|  | 1990 | 233,690 | 425 | - | - | - | - | - | - | - | - | 106,379 | 486 | - | - | 340,069 | 911 |
|  | 1991 | 176,145 | 318 | - | - | - | - | - | - | - | - | 75,084 | 361 | - | - | 251,229 | 679 |
| Faroe Islands | 1982/1983 | 9,086 | - | 101,227 | - | 21,663 | - | 448 | - | 29 | - | - | - | - | - | 132,453 | 625 |
|  | 1983/1984 | 4,791 | - | 107,199 | - | 12,469 | - | 49 | - | - | - | - | - | - | - | 124,508 | 651 |
|  | 1984/1985 | - 324 |  | 123, 510 | - | 9,690 | - | - | - | - | - | - | - | 1,653 | - | 135,776 | 598 |
|  | 1985/1986 | 1,672 |  | 141,740 | - | 4,779 | - | 76 | - | - | - | - | - | 6,287 | - | 154,554 | 545 |
|  | 1986/1987 |  |  | 133,078 | - | 7,070 | - | 80 | - | - | - | - | - | - | - | 140,304 | 539 |
|  | 1987/1988 | 5,833 | - | 55,728 | - | 3,450 | - | 0 | - | - | - | - | - | - | - | 65,011 | 208 |
|  | 1988/1989 | 1,351 | - | 86,417 | - | 5,728 | - | 0 | - | - | - | - | - | - | - | 93,496 | 309 |
|  | 1989/1990 | 1,560 |  | 103,407 | - | 6,463 | - | 6 | - | - | - | - | - | - | - | 111,430 | 364 |
|  | 1990/1991 | 1631 | - | 52,420 | - | 4,390 | - | 8 | - | - | - | - | - | - | - | 57,442 | 202 |
| Finland | $\begin{aligned} & 1990 \\ & 1991 \end{aligned}$ | 13,460 | 24 | - | - | - | - | - | - | - | - | 5,240 | 35 | - | - | 18,700 | 59 |
| France | 1985 | 1,074 | - | - | - | - | - | - | - | - | - | 3,278 | - |  | - - | 4,352 | 22 |
|  | 1986 |  | - | - | - | - | - | - | - | - | - |  | - |  | - - | 6,801 | 28 |
|  | 1987 | 6,013 | 18 | - | - | - | - | - | - | - | - | 1,806 | 9 |  | - - | 7,819 | 27 |
|  | 1988 | 2,063 | 7 | - | - | - | - | - | - | - | - | 4,964 | 25 |  | - - | 7,027 | 32 |
|  | 1989 | 1,351 | 4 | - | - | - | - | - | - | - | - | 1,296 | 6 |  | - - | 2,647 | 10 |
|  | 1990 | 1,886 | 5 | 2,186 | 9 | 146 | 1 | - | - | - | - | - | - |  | - - | 4,218 | 15 |
|  | 1991 | 1,362 | 3 | 1,935 | 6 | 190 | 1 | - | - | - | - | - | - |  | - - | 3,487 | 13 |
| Iceland | 1982 | 23,026 | 58 | - | - | - | - | - | - | - | - | 18,119 | 89 |  | - - | 41,145 | - 147 |
|  | 1983 | 33,769 | 85 | - | - | - | - | - | - | - | - | 24,454 | 113 |  | - - | 58,223 | 198 |
|  | 1984 | 18,901 | 47 | - | - | - | - | - | - | - | - | 22,188 | 112 |  | - - | 41,089 | 159 |
|  | 1985 | 50,000 | 125 | - | - | - | - | - | - | - | - | 16,300 | 94 |  | - - | 66,300 | 217 |
|  | 1986 | 67,300 | 174 | - | - | - | - |  |  | - | - | 22,300 | 136 |  | - - | 89,600 | 310 |
|  | 1987 | 42,550 | 114 | - | - | - | - | - | - | - | - | 18,840 | 108 |  | - | 61,390 | 222 |
|  | 1988 | 112,000 | 288 | - | - | - | - | - | - | - | - | 19,000 | 108 |  | - - | 133,500 | 396 |
|  | 1989 | 70,817 | 158 | - | - | - | - | - | - | - | - | 20,037 | 115 |  | - | 90,854 | 278 |
|  | 1990 | 98,241 | - | - | _ | _ | - | - | - | - | - | 34,267 | - |  | - | 132,508 | 426 |
|  | 1991 | 141,062 | - | - | - | - | - | - | - | - | - | 28,731 | - |  | - - | 169,793 | - 519 |

Table 2.2.1 (cont'd)

| Country | Year | 1SW |  | 2SW |  | 3SW |  | 4SW |  | 5SW |  | MSW! |  | PS |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. | Wt | No. | Wt | No. | Wt | No. |  | No. |  | No. | Wt | No. |  | No. | Wt |
| Ireland | 1980 | 248,333 | 745 | - | - | - | - | - | - | - | - | 39,608 | 202 | - | - | 287,941 | 947 |
|  | 1981 | 173,667 | 521 | - | - | - | - | - | - | - | - | 32,159 | 164 |  | - | 205,826 | 685 |
|  | 1982 | 310,000 | 930 | - | - | - | - | - | - | - | - | 12,353 | 63 | - | - | 322,353 | 993 |
|  | 1983 | 502,000 | 1,506 | - | - | - | - | - | - | - | - | 29,411 | 150 |  | - | 531,411 | 1,656 |
|  | 1984 | 242,666 | 728 | - | - | - | - | - | - | - | - | 19,804 | 101 | - | - | 262,470 | -829 |
|  | 1985 | 498,333 | 1,495 | - | - | - | - | - | - | - | - | 19,608 | 100 | - |  | 517,941 | 1,595 |
|  | 1986 | 498,125 | 1,594 | - | - | - | - | - | - | - | - | 28,335 | 136 | - | - | 526,450 | 1,730 |
|  | 1987 | 358,842 | 1,112 | - | - | - | - | - | - | - | - | 27,609 | 127 | - | - | 386,451 | 1,239 |
|  | 1988 | 559,297 | 1,733 | - | - | - | - | - | - | - | - | 30,599 | 141 | - | - | 589,896 | 1,874 |
|  | 1989 | - |  | - | - | - | - | - | - | - | - | 30,599 |  | - | - | 330,558 | -975 |
|  | 1990 | - | - | - | - | - | - | - | - | - | - | - | - | - |  | 194,785 | 586 |
|  | 1991 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 144,374 | 422 |
| Norway | 1981 | 221,566 | 467 | - | - | - | - | - | - | - | - | 213,943 | 1,189 | - | - | 435,509 | 1,656 |
|  | 1982 | 163,120 | 363 | - | - | - | - | - | - | - | - | 174,229 | 985 | - |  | 337,349 | 1,348 |
|  | 1983 | 278,061 | 593 | - | - | - | - | - | - | - | - | 171,361 | 957 | - |  | 449,442 | 1,550 |
|  | 1984 | 294,365 | 628 | - | - | - | - | - | - | - | - | 176,716 | 995 | - | - | 471,081 | 1,623 |
|  | 1985 | 299,037 | 638 | - | - | - | - | - | - | - | - | 162,403 | 923 | - | - | 461,440 | 1,561 |
|  | 1986 | 264,849 | 556 | - | - | - | - | - | - | - | - | 191,524 | 1,042 | - |  | 456,373 | 1,598 |
|  | 1987 | 235,703 | 491 | - | - | - | - | - | - | - | - | 153,554 | 894 |  | - | 389,257 | 1,385 |
|  | 1988 | 217,617 | 420 | - | - | - | - | - | - | - | - | 120,367 | 656 |  | - | 337,984 | 1,076 |
|  | 1989 | 220,170 | 436 | - | - | - | - | - | - | - | - | 80,880 | 469 |  |  | 301,050 | , 905 |
|  | 1990 | 192,500 | 385 | - | - | - | - | - | - | - | - | 91,437 | 545 |  | - | 286,466 | 930 |
|  | 1991 | 173,500 | 347 | - | - | - | - | - | - | - | - | 92,759 | 538 | - | - | 266,259 | 885 |
| Russia | 1987 | 97,242 | - | 27,135 | - | 9,539 | - | 556 | - | 18 | - | - | - | 2,521 | - | 139,011 | 564 |
|  | 1988 | 53,158 | - | 33,395 | - | 10,256 | - | 294 | - | 25 | - | - |  | 2,937 | - | 100,065 | 419 |
|  | 1989 | 78,023 | - | 23,123 | - | 4,118 | - | 26 | - | - | - | - |  | 2,187 | - | 107,477 | 359 |
|  | 1990 | 70,595 | - | 20,633 | - | 2,919 | - | 101 | - | - | - | - | - | 2,010 | - | 96,258 | 315 |
|  | 1991 | 40,603 | - | 12,458 | - | 3,060 | - | 650 | - | - | - | - | - | 1,375 | - | 58,146 | 215 |
| UK <br> (England <br> \& Wales) | 1985 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 95,531 | 361 |
|  | 1986 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 110,794 | 430 |
|  | 1987 | 66,371 | - | - | - | - | - | - | - | - | - | 17,063 | - | - | - | 83,434 | 302 |
|  | 1988 | 76,521 | - | - | - | - | - | - | - | - | - | 33,642 | - | - | - | 110,163 | 395 |
|  | 1989 | 65,450 | - | - | - | - | - | - | - | - | - | 19,550 | - | - | - | 85,000 | 296 |
|  | 1990 | 53,143 | - | - | - | - | - | - | - | - | - | 33,533 | - | - | - | 86,676 | 338 |
|  | 1991 | 33,000 | - | - | - | - | - | - | - | - | - | 18,000 | - | - | - | 51,000 | 199 |

Table 2.2.1 (cont'd)

| Country | Year | 1SW |  | 2SW |  | 3SW |  | 4SW | 5SW | MSW! |  | PS |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. | Wt | No. | Wt | No. | Wt | No. Wt | No. Wt | No. | Wt | No. |  | No. | Wt |
| UK (Scotland) | 1982 | 208,061 | 416 | - | - | - | - | - - | - | 128,242 | 596 | - |  | 336,303 | 1,092 |
|  | 1983 | 209,617 | 549 | - | - | - | - | - - | - - | 145,961 | 672 | - | - | 320,578 | 1,221 |
|  | 1984 | 213,079 | 509 | - | - | - | - | - - | - | 107,213 | 504 | - | - | 230,292 | 1,013 |
|  | 1985 | 158,012 | 399 | - | - | - | - | - - | - | 114,648 | 514 | - | - | 272,660 | 913 |
|  | 1986 | 202,861 | 526 | - | - | - | - | - - | - - | 148,398 | 745 | - | - | 351,259 | 1,271 |
|  | 1987 | 164,785 | 419 | - | - | - | - | - - | - | 103,994 | 503 | - | - | 268,779 | 922 |
|  | 1988 | 149,098 | 381 | - | - | - | - | - - |  | 112,162 | 501 | - | - | 261,260 | 882 |
|  | 1989 | 174,941 | 431 | - | - | - | - | - - |  | 103,886 | 464 |  | - | 278,827 | 895 |
|  | 1990 | 68,135 | 169 | - | - | - | - | - - | - - | 76,650 | 374 |  | - | 144,785 | 543 |
| USA | 1982 | 33 | - | 1,206 | - | 5 | - | - - | - - | - | - | 21 | - | 1,265 | 6.4 |
|  | 1983 | 26 | - | 314 | 1.2 | 2 | - | - - | - - | - | - | 6 | - | 348 | 1.3 |
|  | 1984 | 50 | - | 545 | 2.1 | 2 | - | - - | - - | - | - | 12 | - | 609 | 2.2 |
|  | 1985 | 23 | - | 528 | 2.0 | 2 | - | - - | - - | - | - | 13 | - | 557 | 2.1 |
|  | 1986 | 76 | - | 482 | 1.8 | 2 | - | - - | - - | - | - | 3 | - | 541 | 1.9 |
|  | 1987 | 33 | - | 229 | 1.0 | 10 | - | - - | - - | - | - | 10 | - | 282 | 1.2 |
|  | 1988 | 49 | - | 203 | 0.8 | 3 | - | - - | - - | - | - | 4 | - | 259 | 0.9 |
|  | 1989 | 157 | 0.3 | 325 | 1.3 | 2 | - | - - | - - | - | - | 3 | - | 487 | 1.6 |
|  | 1990 | 52 | 0.1 | 562 | 2.2 | 12 | - | - - | - - | - | - | 16 | - | 642 | 2.3 |
|  | 1991 | 48 | 0.1 | 185 | 0.7 | 1 | - | - - | - - | - | - | 4 | - | 238 | 0.8 |
| West Greenland | 1982 | 315,532 | - | 17,810 | - | - | - | - - | - - | - | - | 2,688 | - | 336,030 | 1,077 |
|  | 1983 | 90,500 | - | 8,100 | - | - | - | - - |  | - | - | 1,400 | - | 100,000 | 310 |
|  | 1984 | 78,942 | - | 10,442 | - | - | - | - - | - - | - | - | 630 | - | 90,014 | 297 |
|  | 1985 | 292,181 | - | 18,378 | - | - | - | - - | - - | - | - | 934 | - | 311,493 | 864 |
|  | 1986 | 307,800 | - | 9,700 | - | - | - | - - |  | - | - | 2,600 | - | 320,100 | 960 |
|  | 1987 | 297,128 | - | 6,287 | - | - | - | - - | - - | - | - | 2,898 | - | 306, 313 | 966 |
|  | 1988 | 281,356 | - | 4,602 | - | - | - | - - | - - | - | - | 2,296 | - | 288,233 | 893 |
|  | 1989 | 110,359 | - | 5,379 | - | - | - | - |  | - | - | 1,875 | - | 117,613 | 337 |
|  | 1990 | 80,589 | - | 2,772 | - | - | - | - - | - - | - | - | 713 | - | 84,074 | 227 |
|  | 1991 | 155,098 | 384 | 8,144 | 49 | - | - | - - | - | - | - | 706 | 4 | 163,901 | 438 |

!MSW includes all sea ages $>1$, when this cannot be broken down.

Table 2.3.1 Guess-estimates of unreported catches in tonnes within national EEZs in the North-East and North American Commissions of NASCO. Unreported catches for West Greenland Commission are unavailable.

Unreported catches

| Year | North-East | North American | Total |
| :--- | :---: | :---: | ---: |
| $====================================$ |  |  |  |
| 1986 | - | 315 | 315 |
| 1987 | 2,554 | 234 | 2,788 |
| 1988 | 3,087 | 161 | 3,248 |
| 1989 | 2,103 | 174 | 2,277 |
| 1990 | 1,779 | 111 | 1,890 |
| 1991 | 1,555 | 127 | 1,682 |
| $=========================================$ |  |  |  |
| Mean | 2,216 | 187 | 2,377 |

Table 3.2.1.1 Trend analysis of total catches of salmon in home waters for last 30 years.

| Country | Time period years | Average catch tonnes | Coeff. var. \% | Slope | Corr. Coeff. | Prob. level |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Complete Time Series |  |  |  |  |  |
| Finland | 19 | 49 | 29 | - 0.48 | -0.19 | 0.43 |
| France | 19 | 20 | 38 | 0.14 | 0.10 | 0.67 |
| Iceland | 31 | 207 | 38 | 6.28 | 0.73 | $<0.01$ |
| Ireland | 31 | 1,395 | 31 | -5.19 | -0.11 | 0.56 |
| Norway | 31 | 1,546 | 20 | -18.95 | -0.56 | <0.01 |
| Sweden | 31 | 28 | 40 | 0.17 | 0.13 | 0.74 |
| Russia | 31 | 588 | 31 | -9.81 | -0.49 | <0.01 |
| UK (England) | 31 | 358 | 21 | 0.99 | 0.12 | 0.52 |
| UK (Scotland) | 31 | 1,356 | 28 | -31.52 | -0.75 | <0.01 |
| UK (N.Ireland) | 31 | 189 | 53 | $-7.87$ | -0.71 | <0.01 |
| Last Fifteen Years |  |  |  |  |  |  |
| Finland | 15 | 46 | 24 | -0.01 | -0.03 | 0.99 |
| France | 15 | 20 | 36 | 0.43 | 0.26 | 0.35 |
| Iceland | 15 | 249 | 32 | 8.93 | 0.49 | 0.06 |
| Ireland | 15 | 1,231 | 32 | -2.82 | -0.03 | 0.90 |
| Norway | 15 | 1,424 | 21 | -35.58 | -0.52 | 0.04 |
| Russia | 15 | 514 | 25 | -12.20 | -0.42 | 0.11 |
| Sweden | 15 | 29 | 48 | 2.35 | 0.75 | $<0.01$ |
| UK (England) | 15 | 347 | 21 | 3.63 | 0.22 | 0.42 |
| UK (Scotland) | 15 | 1,049 | 18 | -25.90 | -0.63 | 0.01 |
| UK (N.Ireland) | 15 | 114 | 27 | -1.36 | -0.20 | 0.48 |

Table 3.2.1.2. Results of the regression analyses of wild smolt counts with year of migration.

| Country <br> and river | Time <br> period <br> years | Mean | Coeff. Slope <br> var. | Corr. <br> coeff. |
| :--- | :--- | :--- | :--- | :--- |


| IRELAND |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Burrishoole | 22 | 9,328 | 39 | -474 | -0.84 | $<0.01$ |
|  | 10 | 3,804 | 55 | -529 | -0.72 | 0.01 |
| NORWAY |  |  |  |  |  |  |
| Imsa | 10 | 1,152 | 47 | 114 | 0.57 | 0.05 |
| UK (N. Irl) |  |  |  |  |  |  |
| Bush | 10 | 10,694 | 59 | -479 | -0.25 | 0.52 |
| UK (Scot) |  |  |  |  |  |  |
| Girnock Burn | 25 | 2,320 | 29 | -22 | -0.07 | 0.23 |
|  | 10 | 2,006 | 33 | -60 | -0.34 | 0.57 |
| N. Esk | 20 | 175,350 | 29 | -398 | 0.25 | 0.77 |
|  | 7 | 180,143 | 22 | 3,158 | 0.21 | 0.45 |

Table 3.2.2.1. Results of the regressions of adult counts against year.

|  | Complete time series |  |  |  |  |  | Last 10 years |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| River | Time period years | Average count number | Coeff. var. \% | Slope | Corr. coeff. | Prob. level | Average count number | Coeff. var. \% | Slope | Corr. coeff. | Prob. level |
|  |  | ==- |  |  |  | === | = | = $=$ | === | ==== | $=$ = = = |
| England |  |  |  |  |  |  |  |  |  |  |  |
| Severn | 12 | 2,324 | 57 | -98 | -0.27 | . 40 | 2,059 | 63 | 20 | 0.05 | . 90 |
| Ireland |  |  |  |  |  |  |  |  |  |  |  |
| Burrishoole | 22 | 713 | 52 | -48 | -0.70 | . 00 | 486 | 27 | -5 | -0.13 | . 72 |
| N. Ireland |  |  |  |  |  |  |  |  |  |  |  |
| Bush | 19 | 2,181 | 38 | -40 | -0.33 | . 17 | 1,923 | 42 | 92 | 0.34 | . 33 |
| Norway |  |  |  |  |  |  |  |  |  |  |  |
| Imsa | 10 | 41 | 59 | 3 | 0.33 | . 35 | 41 | 59 | 3 | 0.33 | . 35 |
| Russia |  |  |  |  |  |  |  |  |  |  |  |
| Ponoy | 26 | 21,467 | 51 | 525 | 0.37 | . 08 | 24,232 | 39 | 948 | 0.30 | . 40 |
| Russia |  |  |  |  |  |  |  |  |  |  |  |
| Tuloma | 40 | 6,665 | 43 | 57 | 0.23 | . 15 | 8,039 | 28 | 176 | 0.24 | . 51 |
| Scotland |  |  |  |  |  |  |  |  |  |  |  |
| Girnock Burn | 26 | 134 | 52 | -2 | -0.26 | . 19 | 121 | 61 | 4 | 0.17 | . 64 |
| Scotland |  |  |  |  |  |  |  |  |  |  |  |
| North Esk | 11 | 8,472 | 23 | -11 | -0.02 | . 96 | 8,417 | 24 | 22 | 0.03 | . 93 |
| Wales |  |  |  |  |  |  |  |  |  |  |  |
| Dee | 16 | 3,126 | 61 | 6 | 0.01 | . 96 | 3,076 | 71 | 59 | 0.08 | . 82 |
| Sweden |  |  |  |  |  |  |  |  |  |  |  |
| Hogvadsan | 38 | 412 | 91 | 17 | 0.51 | . 01 | 865 | 50 | 19 | 0.14 | . 71 |

Table 3.2.3.1. Regression analysis of postsmolt-survival with year.
Time Average Coeff. Slope Corr. Prob. period postsmolt var. coeff. level
River(1) years survival \% \%
 POSTSMOLT SURVIVAL ENTIRE SERIES

| Burrishoole | 11 | 11.4 | 61 | 0.59 | 0.28 | 0.39 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Ireland $R$

| Bush <br> N.Ireland R | 8 | 11.1 | 39 | -0.46 | -0.27 | 0.52 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Imsa <br> Norway R | 9 | 6.7 | 57 | 0.15 | 0.12 | 0.76 |
| Imsa | 9 | 13.6 | 42 | -0.32 | -0.15 | 0.69 |
| Norway W |  |  | 13.0 | 24 | -0.07 | -0.24 |

Sweden R
RETURN TO RIVER WHOLE SERIES


Table 3.3.1.1. Counts of small salmon from fishways and counting fences in insular Newfoundland 1974-91 by Salmon Fishing Area (SFA); also shown are means, standard deviations (SD), and coefficients of variation (CV).

| Year | Fishways |  |  |  |  |  |  |  |  | Counting fences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SFA 4 |  |  | SFA 5 |  |  | $\frac{\text { SFA } 9}{6}$ | $\frac{\text { SFA } 10}{7}$ | $\frac{\text { SFA } 11}{8}$ | $\frac{\text { SFA 4 }}{9}$ | SFA 9 |  |  | $\frac{\text { SFA } 11}{13}$ |
|  | 1A | 1B | 2 | 3 | 4 | 5 |  |  |  |  | 10 | 11 | 12 |  |
| 1974 | 2538 |  | 857 | 770a |  | 162 |  | 223 |  |  |  |  |  |  |
| 1975 | 9218 | 5531 |  | 1119a |  | 778 |  | 186a |  |  |  |  |  |  |
| 1976 | 3991 | 2935 |  |  |  | 335 |  | 294 |  |  |  |  |  |  |
| 1977 | 6148 | 4300 |  |  |  | 371 |  |  |  |  |  |  |  |  |
| 1978 | 3790 | 2704 | 755 | 1403 | 810 | 436 |  | 390 |  |  |  |  |  |  |
| 1979 | 6715 | 3925 | 404a | 1350a | 569 | 455 |  | 454 |  |  |  |  |  |  |
| 1980 |  | 4597 | 997 | 1712 | 843 | 420 |  | 433 |  |  |  |  |  |  |
| 1981 | 8114 a | 4264 | 2459 | 2414 | 1115 | 619 |  | 334a |  |  |  |  |  |  |
| 1982 | 7605a | 2796 | 1425 | 1281 | 963 | 625 |  | 86a |  |  |  |  | 133 |  |
| 1983 |  | 2952a | 978 | 1195 | 1210 | 853 |  | 233 |  |  | 2330 |  | 272 |  |
| 1984 | 17219 | 6300a | 1081 | 1379 | 1233 | 904 |  | 419 |  |  | 2430 | 89 | 359 |  |
| 1985 | 16652 | 5985 | 1663 | 904 | 1557 | 960 |  | 384 |  |  | 1377a | 124 | 170 |  |
| 1986 | 9697 | 3072 | 1064 | 1036 | 1051 | 726 |  | 725 | 211 |  | 2516 | 158 | 296 | 7515 |
| 1987 | 9014 | 2327 | 493a | 914 | 974 | 570 | 80 | 325a | 155a |  | 1302a | 91 | 368 | 9687 |
| 1988 | 8974 | 3433 | 1562 | 772 | 1737 | 795 | 313 | 543 | 149 |  | 1695 | 97 | 205a | 7118 |
| 1989 | 7192 | 1694 | 596 | 496 | 1138 | 668 | 168 | 706 | 175 | 7743 | 889a | 62 | 441 | 4469 |
| 1990 | 6629 | 1057 | 328a | 745 | 1149 | 410a | 401 | 551 | 208 | 7520 | 1657 | 71 | 307a | 4321 |
| 1991 | 5245 | 1060 | 245 | 562 | 873 | 311a | 211 | 353 | 46 a | 6445 | 394 | 99 | 218 | 2086 |
| 1974-83 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Mean | 5400.0 | 3881.5 | 1245.2 | 1601.0 | 918.3 | 505.4 |  | 337.8 |  |  |  |  |  |  |
| SD | 2434.5 | 1001.2 | 637.1 | 494.9 | 230.3 | 211.3 |  | 101.4 |  |  |  |  |  |  |
| CV | 45.08 | 25.79 | 51.16 | 30.91 | 25.08 | 41.81 |  | 30.02 |  |  |  |  |  |  |
| N | 6 | 8 | 6 | 5 | 6 | 10 |  | 6 |  |  |  |  |  |  |
| 1984-89 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Mean | 11458.0 | 3302.2 | 1193.2 | 916.8 | 1281.7 | 770.5 | 187.0 | 555.4 | 178.3 |  | 2213.7 | 103.5 | 326.8 |  |
| 7197.2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SD | 4326.9 | 1643.5 | 430.9 | 291.9 | 301.3 | 146.3 | 117.7 | 157.8 | 31.1 |  | 451.2 | 33.2 | 101.6 |  |
| 2140.8 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $29.74$ | 37.76 | 49.77 | 36.11 | 31.84 | 23.51 | 18.99 | 62.94 | 28.41 | 17.44 |  | 20.38 | 32.08 | 31.09 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| N | 6 | 5 | 5 | 6 | 6 | 6 | 3 | 5 | 3 |  | 3 | 6 | 5 | 4 |

1 Exploits River
(a) Bishop's Falls
(b) Gt. Rattling Brook

2 Gander River (Salmon Brook)
3 Middle Brook

4 L. Terra Nova River
5 U. Terra Nova River
6 Rocky River
7 Northeast River (Placentia)
8 Grand Bank Brook

9 Gander River
10 Biscay Bay River
11 Northeast Brook (Trepassey)
12 Colinet River
13 Conne River

Partial counts: not included in means

Table 3.3.1.2. Counts of large salmon from fishways and counting fences in insular Newfoundland 1974-91 by Salmon Fishing Area (SFA); also shown are means, standard deviations (SD), and coefficients of variation (CV).

| Year | Fishways |  |  |  |  |  |  |  |  | Counting fences |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SFA 4 |  |  | SFA 5 |  |  | $\frac{\text { SFA } 9}{6}$ | $\frac{\text { SFA } 10}{7}$ | $\frac{\text { SFA } 11}{8}$ | $\frac{\overline{\text { SFA } 4}}{9}$ | SFA 9 |  |  | $\frac{\text { SFA } 11}{13}$ |
|  | 1A | 1B | 2 | 3 | 4 | 5 |  |  |  |  | 10 | 11 | 12 |  |
| 1974 | 411 |  | 9 | 77 a |  | 121 |  | 9 |  |  |  |  |  |  |
| 1975 | 1439 | 505 |  | 9 a |  | 52 |  | 36a |  |  |  |  |  |  |
| 1976 | 460 | 117 |  |  |  | 37 |  | 56 |  |  |  |  |  |  |
| 1977 | 581 | 271 |  |  |  | 262 |  |  |  |  |  |  |  |  |
| 1978 | 303 | 81 | 52 | 16 | 20 | 89 |  | 32 |  |  |  |  |  |  |
| 1979 | 277 | 124 | 6a | 54a | 170 | 30 |  | 37 |  |  |  |  |  |  |
| 1980 |  | 426 | 15 | 91 | 39 | 17 |  | 34 |  |  |  |  |  |  |
| 1981 | 1695a | 514 | 33 | 39 | 90 | 28 |  | 62a |  |  |  |  |  |  |
| 1982 | 181a | 122 | 18 | 20 | 19 | 8 |  | 36a |  |  |  |  | 116 |  |
| 1983 |  | 302 | 12 | 75 | 57 | 76 |  | 22 |  |  | 88 |  | 43 |  |
| 1984 | 529 | 111a | 38 | 57 | 107 | 98 |  | 44 |  |  | 83 | 33 | 97 |  |
| 1985 | 183 | 38 | 26 | 27 | 112 | 60 |  | 0 |  |  | 21a | 41 | 42 |  |
| 1986 | 355 | 174 | 12 | 15 | 140 | 58 |  | 39 | 4 |  | 101 | 30 | 31 | 397 |
| 1987 | 310 | 41 | 9 a | 19 | 56 | 38 | 1 | 16a | 2 a |  | 106a | 30 | 55 | 498 |
| 1988 | 147 | 10 | 24 | 14 | 206 | 45 | 6 | 11 | 2 |  | 61 | 19 | 16a | 418 |
| 1989 | 89 | 14 | 24 | 19 | 142 | 51 | 9 | 15 | 7 | 473 | 104a | 18 | 81 | 319 |
| 1990 | 122 | 15 | 7 a | 13 | 144 | 34a | 17 | 25 | 15 | 508 | 71 | 9 | 50a | 361 |
| 1991 | 99 | 40 | 2 | 14 | 114 | 26a | 16 | 8 | 7 a | 670 | 35 | 13 | 18 | 87 |
| 1974-83 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Mean | 578.5 | 270.0 | 23.2 | 48.2 | 65.8 | 72.0 |  | 31.7 |  |  |  |  |  |  |
| SD | 435.7 | 185.7 | 16.4 | 33.4 | 57.5 | 75.5 |  | 15.7 |  |  |  |  |  |  |
| CV | 75.32 | 68.78 | 70.69 | 69.29 | 87.39 | 104.86 |  | 49.53 |  |  |  |  |  |  |
| N | 6 | 8 | 6 | 5 | 6 | 10 |  | 6 |  |  |  |  |  |  |
| 1984-89 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Mean | 268.8 | 55.4 | 24.8 | 25.2 | 127.2 | 58.3 | 5.3 | 21.8 | 4.3 |  | 81.7 | 28.5 | 61.2 |  |
| 408.0 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| SD | 161.2 | 67.7 | 9.2 | 16.3 | 49.6 | 211 | 4.0 | 18.9 | 2.5 |  | 20.0 | 8.7 | 27.4 |  |
| 73.6 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| CV | 60.31 | 122.20 | 37.10 | 64.68 | 38.99 | 36.19 | 75.47 | 86.70 | 58.14 |  | 24.48 | 30.53 | 44.77 |  |
| 18.04 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| N | 6 | 5 | 5 | 6 | 6 | 6 | 3 | 5 | 3 |  | 3 | 6 | 5 | 4 |

[^0]Table 3.3.1.3. Counts of wild Atlantic salmon at fences (Western Arm Brook) and fishway traps in Salmon Fishing Areas 14, 20, 21, and 23. Numbers in parentheses indicate number of salmon returning to Western Arm Brook before removals for Torrent River transfer.

| Year | SFA 14 |  |  |  | $\text { SFA } 20$ |  | SFA 21 |  | SFA 23 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Western Arm Brook |  | Torrent |  | Liscomb |  | LaHave |  | Saint John |  |
|  | Small | Large | Small | Large | Small | Large | Small | Large | Small | Large |
| 1974 | 299(399) | ) 4 | 38 | 3 |  |  | 29 | 2 | 3,389 | 4,775 |
| 1975 | 393 (631) | ) 1 | 191 | 25 |  |  | 38 | 5 | 5,725 | 6,200 |
| 1976 | 420 (520) | ) 0 | 341 | 47 |  |  | 178 | 23 | 6,797 | 5,511 |
| 1977 | 341 | 3 | 789 | 33 |  |  | 292 | 25 | 3,504 | 7,247 |
| 1978 | 285 | 1 | 971 | 21 |  |  | 275 | 67 | 1,584 | 3,034 |
| 19791 | 1,578 | 0 | 1,984 | 39 | 60 | 0 | 856 | 67 | 6,234 | 1,993 |
| 1980 | 430 | 3 | 792 | 63 | 111 | 0 | 1,637 | 288 | 7,555 | 8,157 |
| 1981 | 447 | 1 | 2,101 | 97 | 76 | 6 | 1,866 | 366 | 4,571 | 2,441 |
| 1982 | 387 | 3 | 2,112 | 523 | 252 | 10 | 799 | 256 | 3,932 | 2,262 |
| 19831 | 1,141 | 4 | 2,007 | 442 | 520 | 15 | 1,129 | 213 | 3,623 | 1,712 |
| 1984 | 117 | 0 | 1,805 | 288 | 606 | 48 | 2,043 | 384 | 7,353 | 7,011 |
| 1985 | 162 | 1 | 1,553 | 30 | 507 | 87 | 1,343 | 638 | 5,331 | 6,391 |
| 1986 | 252 | 0 | 2,815 | 90 | 736 | 117 | 1,579 | 584 | 6,347 | 3,656 |
| 1987 | 378 | 1 | 2,505 | 68 | 1,614 | 88 | 2,529 | 532 | 5,097 | 3,088 |
| 1988 | 102 | 1 | 2,065 | 41 | 477 | 76 | 2,464 | 380 | 8,062 | 1,930 |
| 1989 | 414 | 1 | 1,339 | 54 | 532 | 75 | 2,087 | 511 | 8,417 | 3,854 |
| 1990 | 124 | 0 | 2,296 | 82 | 955 | 44 | 1,861 | 596 | 6,486 | 3,163 |
| 1991 | 233 | 1 | 1,415 | 73 | 586 | 38 | 495 | 236 | 5,415 | 3,639 |
| $\begin{aligned} & \text { Mean } \\ & 1974-83 \end{aligned}$ | 3616 | 2 | 1,133 | 129 | 204 | 6 | 710 | 131 | 4,691 | 4,333 |
| $\begin{aligned} & \text { Mean } \\ & 1986-90 \end{aligned}$ | 0254 | 0.6 | 2,204 | 67 | 863 | 80 | 2,104 | 521 | 6,882 | 3,138 |

Table 3.3.1.4 Counts of Atlantic salmon at fishways 1975-91 in areas Q3 and Q7, Quebec.

| Year | $\frac{07}{\text { de la Trinité }}$ |  | Q3 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Matane |  | Madeleine |  | Mitis |  |
|  | Small | Large | $\overline{\text { Small }}$ | Large | Small | Large | Small | Large |
| 1975 |  |  | 943 | 1487 |  |  | 66 | 159 |
| 1976 |  |  | 1067 | 1393 | 115 | 155 | 90 | 165 |
| 1977 |  |  | 1189 | 1078 | 77 | 70 | 83 | 170 |
| 1978 |  |  | 807 | 2571 | 56 | 374 | 77 | 133 |
| 1979 |  |  | 1540 | 725 | 93 | 57 | 281 | 141 |
| 1980 | 1144 | 156 | 1466 | 2102 | 81 | 79 | 193 | 387 |
| 1981 | 1892 | 367 |  |  | 313 | 146 | 270 | 151 |
| 1982 | 2173 | 828 |  |  | 259 | 317 | 114 | 563 |
| 1983 | 891 | 461 |  |  | 96 | 167 | 46 | 157 |
| 1984 | 1663 | 421 | 876 | 931 | 74 | 392 | 239 | 236 |
| 1985 | 1008 | 519 | 762 | 1003 | 156 | 301 | 181 | 378 |
| 1986 | 1364 | 546 | 2364 | 1397 | 359 | 439 | 636 | 451 |
| 1987 | 1115 | 514 | 1018 | 2290 | 406 | 951 | 225 | 557 |
| 1988 | 1324 | 760 | 692 | 2086 | 499 | 781 | 477 | 314 |
| 1989 | 1744 | 441 | 1218 | 923 | 482 | 926 | 338 | 428 |
| 1990 | 1637 | 460 | 1270 | 1520 | 452 | 932 | 528 | 282 |
| 1991 | 1306 | 496 | 1586 | 1354 | 461 | 671 | 329 | 327 |
| Mean |  |  |  |  |  |  |  |  |
| 75-83 | 1525 | 453 | 1169 | 1559 | 136 | 171 | 136 | 225 |
| 86-90 | 1437 | 544 | 1312 | 1643 | 440 | 806 | 441 | 406 |

Table 3.3.1.5. Estimated numbers of wild returning and spawning Atlantic salmon, egg depositions, ratios of large salmon spawners to returns and fraction of target egg deposition attained in 16 rivers in atlantic Canada. Empty cells mean no prediction available. Bold numbers are target spawners and eggs.

|  | Returns ( $10^{3}$ ) |  |  | Spawners ( $10^{3}$ ) |  | $\begin{aligned} & \text { Eggs } \\ & \left(10^{\circ}\right) \end{aligned}$ | Large Spawners/ <br> Large Returns |  | Eggs/ Target |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Small | Large | Predicted Large | Small | Large |  |  |  | Eggs |

Restigouche River - SFA 15

| TARGET |  |  |  | 2.6 | 12.2 | 71.4 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |
| 1982 | 8.0 | 11.2 |  | 2.0 | 1.8 | 10.9 | 0.16 | 0.15 |
| 1983 | 3.4 | 10.2 | 13.5 | 0.6 | 1.4 | 8.7 | 0.14 | 0.12 |
| 1984 | 10.9 | 7.8 | 11.3 | 1.3 | 3.1 | 18.6 | 0.40 | 0.26 |
| 1985 | 7.0 | 10.0 | 12.2 | 2.5 | 6.3 | 37.4 | 0.63 | 0.52 |
| 1986 | 10.7 | 14.1 | 14.8 | 3.8 | 8.8 | 52.6 | 0.63 | 0.74 |
| 1987 | 10.0 | 10.2 | 21.9 | 3.5 | 5.9 | 35.0 | 0.58 | 0.49 |
| 1988 | 13.5 | 12.7 | 12.9 | 4.7 | 8.2 | 49.3 | 0.65 | 0.69 |
| 1989 | 6.7 | 10.6 |  | 2.3 | 6.2 | 37.1 | 0.58 | 0.52 |
| $1990^{2}$ | $10.2-$ | $10.5-$ |  | $4.3-$ | $6.3-$ | $37.9-$ | $0.65^{5}$ | $0.53-$ |
|  | 17.1 | 16.4 |  | 10.1 | 11.3 | 68.0 |  | 0.95 |
| $1991^{2}$ | $5.9-$ | $8.6-$ |  | $2.5-$ | $5.1-$ | $30.4-$ | $0.64^{5}$ | $0.43-$ |
|  | 9.8 | 13.6 |  | 5.9 | 9.3 | 55.5 |  | 0.78 |

Miramichi River ${ }^{1}$-SFA 16

| TARGET | 22.6 | 23.6 | 132.0 |
| :--- | :--- | :--- | :--- |


| 1982 | 80.4 | 30.8 |  | 52.0 | 12.3 | 109.8 | 0.40 | 0.83 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1983 | 25.2 | 27.9 | 43.0 | 10.8 | 7.5 | 48.1 | 0.27 | 0.36 |
| 1984 | 29.7 | 15.1 | 10.2 | 14.9 | 13.7 | 77.0 | 0.91 | 0.58 |
| 1985 | 60.8 | 20.7 | 18.4 | 37.8 | 19.1 | 130.0 | 0.92 | 0.98 |
| 1986 | 117.5 | 31.3 | 28.4 | 85.4 | 29.2 | 226.4 | 0.93 | 1.72 |
| 1987 | 84.8 | 19.4 | 54.2 | 58.8 | 17.1 | 175.9 | 0.88 | 1.33 |
| 1988 | 121.9 | 21.7 | 36.4 | 86.3 | 20.0 | 189.3 | 0.92 | 1.43 |
| 1989 | 75.2 | 17.2 | - | 44.4 | 14.6 | 124.1 | 0.84 | 0.94 |
| 1990 | 83.4 | 28.6 | - | 59.5 | 26.6 | 191.2 | 0.93 | 1.45 |
| 1991 | 60.9 | 29.9 | 26.0 | 45.6 | 28.1 | 200.6 | 0.94 | 1.52 |

Saint John River ${ }^{1}$ above Mactaquac Dam - SFA 23

| TARGET |  |  |  | 3.2 | 4.4 | 29.5 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |  |  |  |
| 1982 | 7.8 | 6.5 |  | 4.9 | 2.3 | 16.5 | 0.35 | 0.56 |
| 1983 | 5.8 | 4.0 |  | 3.7 | 1.3 | 8.5 | 0.33 | 0.29 |
| 1984 | 9.8 | 10.9 | 6.2 | 7.2 | 7.2 | 39.5 | 0.66 | 1.34 |
| 1985 | 8.5 | 11.3 | 10.5 | 4.5 | 6.3 | 36.3 | 0.56 | 1.23 |
| 1986 | 8.8 | 6.9 | 8.8 | 5.9 | 3.5 | 26.1 | 0.51 | 0.88 |
| 1987 | 9.2 | 4.8 | 11.0 | 7.0 | 2.8 | 19.7 | 0.58 | 0.67 |
| 1988 | 10.2 | 3.5 | 8.0 | 7.8 | 1.7 | 12.9 | 0.48 | 0.44 |
| 1989 | 10.9 | 4.5 | 7.1 | 7.5 | 3.5 | 24.7 | 0.77 | 0.84 |
| 1990 | 8.8 | 4.1 |  | 6.1 | 3.2 | 22.4 | 0.78 | 0.76 |
| 1991 | 8.8 | 5.3 |  | 5.7 | 3.5 | 24.2 | 0.66 | 0.82 |

Table 3.3.1.5. (Continued)

| Year | Returns ( $10^{3}$ ) |  |  | Spawners ( $10^{3}$ ) |  | $\begin{aligned} & \text { Eggs } \\ & \left(10^{6}\right) \end{aligned}$ | Large Spawners/ Large Returns |  | Eggs/ <br> Target <br> Eggs |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Small | Large | Predicted Large | Small | Large |  |  |  |  |
| LaHave River above Morgan Falls - SFA 21 |  |  |  |  |  |  |  |  |  |
|  | TARGE |  |  |  | -- | -- | -- |  |  |
|  | 1983 | 1.1 | 0.2 |  | 1.1 | 0.2 | 2.0 | 1.00 |  |
|  | 1984 | 2.0 | 0.4 | 0.23 | 2.0 | 0.3 | 3.1 | 0.75 |  |
|  | 1985 | 1.3 | 0.6 | $0.3^{3}$ | 1.3 | 0.4 | 3.4 | 0.67 |  |
|  | 1986 | 1.6 | 0.6 | 0.43 | 1.6 | 0.4 | 4.1 | 0.67 |  |
|  | 1987 | 2.5 | 0.5 | 0.43 | 2.5 | 0.4 | 4.9 | 0.80 |  |
|  | 1988 | 2.5 | 0.4 | $0.7{ }^{3}$ | 2.4 | 0.3 | 4.4 | 0.75 |  |
|  | 1989 | 2.1 | 0.5 |  | 2.1 | 0.4 | 4.3 | 0.80 |  |
|  | 1990 | 1.9 | 0.4 |  | 1.9 | 0.3 | 3.3 | 0.75 |  |
|  | 1991 | 0.5 | 0.2 |  | 0.4 | 0.2 | 1.4 | 0.73 |  |
| Margaree River ${ }^{1}$ - SFA 18 |  |  |  |  |  |  |  |  |  |
|  | TARGE |  |  |  | 0.6 | 1.0 | 6.7 |  |  |
|  | 1983 | 0.2 | 0.5 |  | 0.1 | 0.3 | 1.8 | 0.60 | 0.27 |
|  | 1984 | 0.4 | 0.4 |  | 0.2 | 0.3 | 2.0 | 0.75 | 0.30 |
|  | 1985 | 0.6 | 0.8 |  | 0.4 | 0.8 | 5.3 | 1.00 | 0.79 |
|  | 1986 | 0.8 | 2.0 |  | 0.5 | 2.0 | 12.9 | 1.00 | 1.93 |
|  | 1987 | 1.5 | 4.0 |  | 1.1 | 4.0 | 25.9 | 1.00 | 3.87 |
|  | 1988 | 2.2 | 1.7 |  | 1.6 | 1.7 | 11.1 | 1.00 | 1.65 |
|  | 1989 | 0.8 | 2.3 |  | 0.6 | 2.3 | 16.9 | 1.00 | 2.17 |
|  | 1990 | 1.0 | 11.1 |  | 0.7 | 11.1 | 71.7 | 1.00 | 10.7 |
|  | 1991 | 1.9 | 3.5 |  | 1.5 | 3.5 | 21.9 | 1.00 | 3.3 |
| Conne River - SFA 11 |  |  |  |  |  |  |  |  |  |
|  | TARGET |  |  |  | 4.0 | - | 7.8 |  |  |
|  | 1986 | 8.3 | 0.4 |  | 5.4 | 0.4 | 11.3 | $0.65{ }^{4}$ | 1.45 |
|  | 1987 | 10.2 | 0.5 |  | 7.8 | 0.5 | 16.7 | $0.77{ }^{4}$ | 2.14 |
|  | 1988 | 7.6 | 0.4 | 7.9.8.8 | 5.6 | 0.4 | 12.4 | $0.74{ }^{4}$ | 1.59 |
|  | 1989 | 5.0 | 0.3 | 6.2-6.8 | 3.6 | 0.3 | 8.0 | $0.72{ }^{4}$ | 1.03 |
|  | 1990 | 5.4 | 0.4 | 6.8-7.9 | 3.8 | 0.4 | 8.7 | $0.70^{4}$ | 1.12 |
|  | 1991 | 2.4 | 0.1 | 4.5-5.3 | 2.1 | 0.1 | 4.0 | $0.88{ }^{4}$ | 0.51 |

Table 3.3.1.5. (Continued)

| Year | Returns ( $10^{3}$ ) |  |  | Spawners ( $10^{3}$ ) |  | $\begin{aligned} & \text { Eggs } \\ & \left(10^{6}\right) \end{aligned}$ | Large Spawners/ Large Returns |  | Eggs/ <br> Target <br> Eggs |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Small | Large | Predicted Large |  | Large |  |  |  |  |
| Rivière de la Trinité - SFA Q7 |  |  |  |  |  |  |  |  |  |
|  | TARGET |  |  |  | 1.0 | 0.5 | 2.7 |  |  |
|  | 1982 | 2.4 | 0.3 |  | 1.6 | 0.2 | 1.2 | 0.66 | 0.44 |
|  | 1983 | 0.9 | 0.5 |  | 0.7 | 0.5 | 2.5 | 1.00 | 0.93 |
|  | 1984 | 1.8 | 0.5 |  | 1.4 | 0.4 | 2.2 | 0.80 | 0.81 |
|  | 1985 | 1.1 | 0.6 |  | 0.9 | 0.4 | 2.2 | 0.67 | 0.81 |
|  | 1986 | 1.6 | 0.6 |  | 1.1 | 0.4 | 2.3 | 0.67 | 0.85 |
|  | 1987 | 1.3 | 0.6 |  | 0.8 | 0.4 | 2.6 | 0.67 | 0.96 |
|  | 1988 | 1.6 | 0.8 |  | 1.0 | 0.7 | 4.5 | 0.88 | 1.67 |
|  | 1989 | 1.9 | 0.5 |  | 1.3 | 0.3 | 2.3 | 0.60 | 0.85 |
|  | 1990 | 1.9 | 0.5 |  | 1.2 | 0.4 | 2.6 | 0.80 | 0.96 |
|  | 1991 | 1.3 | 0.5 |  | 1.0 | 0.4 | 2.4 | 0.77 | 1.04 |
| Humber River - SFA 13 |  |  |  |  |  |  |  |  |  |
| TARGET |  |  |  | 18 | 6 | 27.7 |  |  |  |
|  | 1987 | 12.3 | 0.9 |  | 9.2 | 0.9 | 16.1 |  | 0.58 |
|  | 1988 | 16.2 | 1.1 |  | 12.1 | 1.1 | 12.4 |  | 0.77 |
|  | 1989 | 4.9 | 0.3 |  | 3.7 | 0.3 | 2.9 |  | 0.23 |
|  | 1990 | 12.2 | 0.9 |  | 9.2 | 0.9 | 16.0 |  | 0.58 |
|  | 1991 | 5.7 | 0.4 |  | 4.3 | 0.4 | 7.5 |  | 0.27 |
| Gander River - SFA 4 |  |  |  |  |  |  |  |  |  |
| TARGET |  |  |  | 22 | - | 46.2 |  |  |  |
|  | 1989 | 7.7 | 0.45 |  | 6.6 | 0.45 | 16.3 |  | 0.35 |
|  | 1990 | 7.7 | 0.51 |  | 6.6 | 0.51 | 16.5 |  | 0.36 |
|  | 1991 | 6.7 | 0.67 |  | 5.6 | 0.67 | 15.1 |  | 0.33 |
| Rocky River - SFA 9 |  |  |  |  |  |  |  |  |  |
| TARGET |  |  |  | 0.9 | - | 3.4 |  |  |  |
|  | 1987 | 0.08 |  |  | $0.2{ }^{9}$ |  | 0.8 |  | 0.23 |
|  | 1988 | 0.3 |  |  | 0.3 |  | 1.2 |  | 0.36 |
|  | 1989 | 0.2 |  |  | 0.2 |  | 0.7 |  | 0.20 |
|  | 1990 | 0.4 |  |  | 0.4 |  | 1.6 |  | 0.47 |
|  | 1991 | 0.2 |  |  | 0.2 |  | 0.9 |  | 0.26 |

Table 3.3.1.5. (Continued)

| Year | Returns ( $10^{3}$ ) |  | Spawners (10 ${ }^{3}$ ) |  | $\begin{aligned} & \text { Eggs } \\ & \left(10^{6}\right) \end{aligned}$ | Large Spawners/ <br> Large Returns | Eggs/ Target Eggs |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Large | Predicted Large |  | Large |  |  |  |
| Terra Nova River - SFA 5 |  |  |  |  |  |  |  |
| TARGET ${ }^{8}$ |  |  |  | 7.1 | - | 14.3 |  |
| 1986 | 1.5 | 0.14 |  | 1.0 | 0.14 | 2.7 | 0.19 |
| 1987 | 1.4 | 0.06 |  | 0.9 | 0.06 | 2.2 | 0.15 |
| 1988 | 2.1 | 0.21 |  | 1.7 | 0.21 | 4.3 | 0.30 |
| 1989 | 1.4 | 0.14 |  | 1.1 | 0.14 | 2.9 | 0.20 |
| 1990 | 1.5 | 0.14 |  | 1.1 | 0.14 | 2.9 | 0.20 |
| 1991 | 1.1 | 0.11 |  | 0.8 | 0.11 | 2.2 | 0.16 |
| Middle Brook - SFA 5 |  |  |  |  |  |  |  |
| TARGET |  |  | 1.01 | - | 2.34 |  |  |
| 1986 | 1.0 | 0.015 |  | 0.76 | 0.015 | 2.10 | 0.90 |
| 1987 | 1.1 | 0.019 |  | 0.87 | 0.019 | 2.11 | 0.90 |
| 1988 | 1.3 | 0.014 |  | 0.63 | 0.014 | 1.54 | 0.66 |
| 1989 | 0.6 | 0.019 |  | 0.46 | 0.019 | 1.18 | 0.50 |
| 1990 | 1.1 | 0.013 |  | 0.72 | 0.013 | 1.74 | 0.74 |
| 1991 | 0.8 | 0.014 |  | 0.49 | 0.014 | 1.20 | 0.51 |
| Biscay Bay River - SFA 9 |  |  |  |  |  |  |  |
| TARGET |  |  | 1.11 | - | 2.95 |  |  |
| 1986 | 2.7 | 0.10 |  | 2.18 | 0.10 | 6.14 | 2.08 |
| 1987 | 1.4 | 0.11 |  | 1.17 | 0.11 | 3.52 | 1.19 |
| 1988 | 1.8 | 0.06 |  | 1.33 | 0.06 | 3.74 | 1.26 |
| 1989 | 1.0 | 0.10 |  | 0.81 | 0.10 | 2.56 | 0.87 |
| 1990 | 1.7 | 0.07 |  | 1.32 | 0.07 | 3.78 | 1.28 |
| 1991 | 0.4 | 0.04 |  | 0.40 | 0.04 | 1.16 | 0.39 |
| Grand River ${ }^{10}$ (above fishway) - SFA 19 |  |  |  |  |  |  |  |
| TARGET ${ }^{8}$ |  |  |  | 0.54 | - | 1.1 |  |
| 1988 | 0.70 | - |  | 0.74 | - | - | $1.36{ }^{4}$ |
| 1989 | 0.60 | - |  | 0.45 | - | - | $0.83{ }^{4}$ |
| 1990 | 0.62 | - |  | 0.44 | - | - | $0.83{ }^{4}$ |
| 1991 | 0.44 | - |  | 0.35 | - | - | $0.64{ }^{4}$ |
| Liscomb River - SFA 19 |  |  |  |  |  |  |  |
| TARGET ${ }^{\prime}$ |  |  |  | - | - | - |  |
| 1985 | 0.8 | 0.14 |  | 0.7 | 0.1 | 1.74 |  |
| 1986 | 1.7 | 0.22 |  | 1.5 | 0.2 | 3.52 |  |
| 1987 | 2.4 | 0.14 |  | 2.1 | 0.1 | 4.25 |  |
| 1988 | 1.0 | 0.12 |  | 0.9 | 0.1 | 2.06 |  |
| 1989 | 0.9 | 0.15 |  | 0.8 | 0.1 | 2.02 |  |
| 1990 | 1.6 | 0.07 |  | 1.4 | 0.05 | 2.67 |  |
| 1991 | 0.8 | 0.06 |  | 0.7 | 0.04 | 1.58 |  |

Table 3.3.1.5. (Continued)

| Year | Returns ( $10^{3}$ ) |  |  | Spawners ( $10^{3}$ ) |  | $\begin{aligned} & \text { Egqs } \\ & \left(10^{6}\right) \end{aligned}$ | Large Spawners/ Large Returns | Eggs/ <br> Target <br> Eggs |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Small | Large | Predicted Large | Small | Large |  |  |  |

Northeast River (Placentia Bay) - SFA 10

| TARGET |  |  | 0.22 | - | 0.72 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1986 | 0.88 | 0.039 |  | 0.65 | 0.039 | 2.49 | 3.46 |
| 1987 | 0.35 | 0.016 |  | 0.32 | 0.016 | 1.09 | 1.52 |
| 1988 | 0.64 | 0.011 | 0.45 | 0.011 | 1.50 | 2.09 |  |
| 1989 | 0.81 | 0.015 | 0.60 | 0.015 | 2.00 | 2.77 |  |
| 1990 | 0.70 | 0.025 | 0.53 | 0.025 | 1.81 | 2.51 |  |
| 1991 | 0.37 | 0.008 | 0.35 | 0.008 | 1.16 | 1.61 |  |

[^1]Table 4.1.2.1 Nominal landings of Atlantic salmon by
Faroes vessels in years 1982-1990 and the seasons 1981/1982-1990/1991.

| Year | Catch (t) | Season | Catch ( $t$ ) |
| :--- | :---: | :---: | :---: |
| 1982 | 606 | $1981 / 1982$ | 796 |
| 1983 | 678 | $1982 / 1983$ | 625 |
| 1984 | 628 | $1983 / 1984$ | 651 |
| 1985 | 566 | $1984 / 1985$ | 598 |
| 1986 | 530 | $1985 / 1986$ | 545 |
| 1987 | 576 | $1986 / 1987$ | 539 |
| 1988 | 243 | $1987 / 1988$ | 208 |
| 1989 | 364 | $1988 / 1989$ | 309 |
| 1990 | 315 | $1989 / 19901$ | 364 |
| $1991^{1}$ | 95 | $1990 / 1991$ | 202 |

${ }^{1}$ Preliminary catch.

Table 4.1.2.2 Catch in number of salmon by month in the Faroes fishery for the seasons 1983/1984 to 1990/1991.

| Season | Nov | Dec | Jan | Feb | Mar | Apr | May | Jun | Total |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $1983 / 1984$ | 8,680 | 24,882 | 12,504 | 26,396 | 32,712 | 12,486 | 6,849 | 0 | 124,508 |
| $1984 / 1985$ | 5,884 | 20,419 | 14,493 | 24,380 | 26,035 | 25,471 | 19,095 | 0 | 135,776 |
| $1985 / 1986$ | 1,571 | 27,611 | 13,992 | 50,146 | 25,968 | 21,209 | 14,057 | 0 | 154,554 |
| $1986 / 1987$ | 1,881 | 19,693 | 5,905 | 15,113 | 35,241 | 21,953 | 39,153 | 1,365 | 140,304 |
| $1987 / 1988$ | 4,259 | 27,125 | 5,803 | 9,387 | 9,592 | 4,203 | 4,642 | 0 | 65,011 |
| $1988 / 1989$ | 17,019 | 24,743 | 2,916 | 4,663 | 12,457 | 31,698 | - | - | 93,496 |
| $1989 / 1990$ | 13,079 | 40,168 | 5,533 | 11,282 | 11,379 | 29,504 | 570 | - | 111,425 |
| $1990 / 1991$ | 6,921 | 28,972 | 3,720 | 7,996 | 6,275 | 3,557 | - | - | 57,442 |

Table 4.1.2.3 Estimation of discard rates in the Faroes fishery 1982/1983 to 1990/1991.

| Season | No. of <br> samples | Number <br> sampled | No. <br> < 60 cm | Discard <br> rate \% | Range \% |  |
| :--- | :---: | ---: | ---: | ---: | ---: | ---: | ---: |

Table 4.1.3.1 Catch of salmon in number per unit effort (1,000 hooks) by month in the Faroes longline fishery south of $65^{\circ} 30^{\prime} \mathrm{N}$ in the seasons 1981/1982-1990/1991.

| Season | Nov | Dec | Jan | Feb | Mar | Apr | May | Jun | Season |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $1981 / 1982$ | - | 38 | 41 | 49 | 58 | 51 | 34 | - | 46 |
| $1982 / 1983$ | 19 | 120 | - | 61 | 50 | 39 | 36 | 40 | 48 |
| $1983 / 1984$ | 85 | 80 | 86 | 58 | 45 | 28 | 26 | - | 51 |
| $1984 / 1985$ | 38 | 38 | 32 | 32 | 37 | 39 | 40 | - | 36 |
| $1985 / 1986$ | 64 | 52 | 68 | 54 | 48 | 78 | 61 | - | 56 |
| $1986 / 1987$ | 31 | 43 | 34 | 44 | 70 | 111 | 102 | - | 64 |
| $1987 / 1988$ | 56 | 51 | - | 47 | 34 | 25 | 22 | - | 43 |
| $1988 / 1989$ | 63 | 80 | 48 | 68 | 61 | 76 | - | - | 71 |
| $1989 / 1990$ | 81 | 86 | 38 | 56 | 87 | 77 | - | - | 76 |
| $1990 / 1991$ | 81 | 97 | - | 35 | 39 | 51 | - | - | 67 |

Table 4.1.3.2 Catch of salmon in number per unit effort (1,000 hooks) by month in the Faroes longline fishery north of $65^{\circ} 30^{\prime} \mathrm{N}$ in the seasons 1981/1982-1990/1991.

| Season | Nov | Dec | Jan | Feb | Mar | Apr | May | Jun | Season |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $1981 / 1982$ | - |  | 72 | 69 | 73 | 64 | 65 | - | 69 |
| $1982 / 1983$ | - | - | - | - | 68 | 41 | - | $54^{1}$ | 60 |
| $1983 / 1984$ | $102^{1}$ | - | - | - | $34^{1}$ | - | - | - | 70 |
| $1984 / 1985$ | - | - | - | 46 | 31 | 37 | 43 | - | 37 |
| $1985 / 1986$ | - | - | - | - | 38 | 82 | 84 | - | 80 |
| $1986 / 1987$ | - | - | $67^{1}$ | 64 | 77 | - | 94 | - | 77 |
| $1987 / 1988$ | 48 | 68 | 73 | $71^{1}$ | 31 | $32^{1}$ | - | - | 65 |
| $1988 / 1989$ | - | - | - | - | $71^{1}$ | - | - | - | 71 |
| $1989 / 1990$ | - | - | - | - | - | 103 | - | - | 103 |
| $1990 / 1991$ | - | - | - | - | - | - | - | - | - |

[^2]Table 4.1.4.1 Percentage sea age distribution of samples by month in the 1990/1991 season determined by fork length method, see text for details.

|  | Sea age |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
| Month | 1 | 2 | $3+$ | Tota1 |
| Nov/Dec | 1.4 | 96.4 | 2.2 | 100.0 |
| Jan | 0 | 88.4 | 11.6 | 100.0 |
| Feb | 1.5 | 81.5 | 17.0 | 100.0 |
| Mar | 0 | 77.5 | 22.5 | 100.0 |
| Apr | 0.5 | 88.5 | 11.0 | 100.0 |
| Weighted |  |  |  |  |
| mean | 1.0 | 90.7 | 8.3 | 100.0 |

Table 4.1.4.2 Catch in number by sea age class by fishing seasons in the Faroes salmon fishery since 1983/1984.

|  | Sea Age |  |  |  |  |  |  |  |  |  | Total |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
| Season | 1 | $\%$ | 2 | $\%$ | 3 | $\%$ | 4 | $\%$ |  |  |  |
| $1983 / 1984$ | 5,142 | 3 | 136,418 | 86 | 16,401 | 10 | 59 | 0 | 157,961 |  |  |
| $1984 / 985$ | 3811 | 0 | 138,375 | 92 | 11,358 | 8 | 0 | 0 | 150,114 |  |  |
| $1985 / 1986$ | 2,021 | 1 | 169,462 | 96 | 5,671 | 3 | 87 | 0 | 177,241 |  |  |
| $1986 / 1987$ | 71 | 0 | 124,628 | 95 | 6,621 | 5 | 75 | 0 | 131,395 |  |  |
| $1987 / 1988$ | 5,833 | 9 | 55,728 | 86 | 3,450 | 5 | 0 | 0 | 65,011 |  |  |
| $1988 / 1989$ | 1,351 | 1 | 86,417 | 92 | 5,728 | 6 | 0 | 0 | 93,496 |  |  |
| $1989 / 1990$ | 1,560 | 1 | 103,407 | 93 | 6,463 | 6 | 0 | 0 | 111,430 |  |  |
| $1990 / 1991$ | 631 | 1 | 52,420 | 91 | 4,390 | 8 | 0 | 0 | 57,442 |  |  |
| Tota1 | 16,990 | 2 | 866,855 | 92 | 60,082 | 6 | 221 | 0 | 944,148 |  |  |

Table 4.1.5.1 Number of microtags recovered at Faroes from European countries.

| Season | Country of origin | Discards <br> Recovery | 1SW1 | 2SW | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1981/1982 | Ireland | 1 | - | 2 | 3 |
|  | UK (Scotland) | - | - | 2 | 2 |
| 1982/1983 | Ireland | 4 | 2 | 2 | 8 |
|  | UK (Scotland) | - | - | 1 | 1 |
| 1983/1984 | UK (Scotland) | - | - | 1 | 1 |
| 1984/1985 | Iceland | 2 | - | - | 2 |
|  | Ireland | 15 | - | 3 | 18 |
|  | UK (Scotland) | 3 | - | - | 3 |
|  | Raising Factors | 16.4 | 3.55 | 3.55 |  |
| 1985/1986 | Ireland | 8 | - | 5 | 13 |
|  | Faroe Islands | - | - | 3 | 3 |
|  | UK (England and Wales) | - | - | 1 | 1 |
|  | Raising Factors | 10.7 | 3 | 3 |  |
| 1986/1987 | Faroe Islands | - | - | 29 | 29 |
|  | Ireland | 8 | - | 1 | 9 |
|  | UK (England and Wales) | 1 | - | 5 | 6 |
|  | UK (N. Ireland) | 4 | - |  | 4 |
|  | UK (Scotland) | 2 | - | 1 | 3 |
|  | Raising Factors | 3.8 | 3 | 3 |  |
| 1987/1988 | Faroe Islands | - | - | 20 | 20 |
|  | Iceland | - | 1 | - | 1 |
|  | Ireland | 3 | 1 | 4 | 8 |
|  | UK (England and Wales) | 1 | - | 3 | 4 |
|  | Raising Factors | 51.4 | 2.7 | 2.7 |  |
| 1988/1989 | Canada | 1 | - | - | 1 |
|  | Faroe Islands | 2 | - | - | 2 |
|  | Iceland | - | - | 15 | 15 |
|  | Ireland | 17 | - | 2 | 19 |
|  | UK (England and Wales) | 2 | 1 | 13 | 16 |
|  | UK (N. Ireland) | - | - | 1 | 1 |
|  | UK (Scotland) | 2 | - | 2 | 4 |
|  | USA | - | - | 1 | 1 |
|  | Raising Factors | 6.1 | 1.8 | 1.8 |  |
| 1989/1990 | Faroe Islands | - | - | 30 | 30 |
|  | Ireland | 14 | - | 3 | 17 |
|  | UK (England and Wales) | 3 | 1 | 5 | 9 |
|  | Raising Factors | 7.5 | 2.3 | 2.3 |  |
| 1990/1991 | Faroe Islands | - | - | 2 | 2 |
|  | Iceland | - | - | 1 | 1 |
|  | Ireland | 3 | - | - | 3 |
|  | UK (England and Wales) | 1 | - | 4 | 5 |
|  | UK (N. Ireland) | 1 | - | - | 1 |
|  | UK (Scotland) | 1 | - | 1 | 2 |
|  | Raising Factors | 14.6 | 3.8 | 3.8 |  |

Table 4.1.5.2 Calculation of the raising factors for the microtag data from the Faroes fishery 1984/85 to 1990/91

|  |  | A | B | C | D | E | F | $\begin{gathered} \mathrm{H} \\ \text { 1SW AND 2SW } \\ \text { RAISING } \end{gathered}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | DISCARD |  |  |
|  |  |  | DISCARD |  |  | TOTAL | RAISING |  |  |
| YEAR OF | No. | TOTAL | NO. OF | RATE \% | TOTAL | DISCARD | FACTOR | TOTAL | FACTOR |
| FISHERY | TRIPS | SAMPLE | DISCARD | $B / A * 100$ | LANDED | C*D/(100-C) | E/B | OBSERVED | D/G |
| 1984/85 | 12 | 9,546 | 1,289 | 13.5 | 135,776 | 21,196 | 16.4 | 38,276 | 3.55 |
| 1985/86 | 7 | 14,654 | 268 | 1.8 | 154,554 | 2,879 | 10.7 | 52,186 | 2.96 |
| 1986/87 | 13 | 39,758 | 2,849 | 7.2 | 140,304 | 10,830 | 3.8 | 47,347 | 2.96 |
| 1987/88 | 2 | 1,499 | 235 | 15.7 | 65,011 | 12,087 | 51.4 | 24,160 | 2.69 |
| 1988/89 | 9 | 17,235 | 1,804 | 10.5 | 93,496 | 10,930 | 6.1 | 51,562 | 1.81 |
| 1989/90 | 5 | 16,375 | 1,533 | 9.4 | 111,425 | 11,509 | 7.5 | 48,352 | 2.30 |
| 1990/91 | 3 | 4,615 | 681 | 14.8 | 57,442 | 9,944 | 14.6 | 14,902 | 3.85 |

Table 4.1.5.3 Estimated numbers of discards, 1 SW and 2 SW microtagged salmon caught in the Faroese fishery from smolt releases between 1984 and 1990 (year of fishery for 2 SW is $\mathrm{n}+1$ ).

| Year of migration yr(n) | Country <br> of origin | Number released | $\begin{gathered} \text { Discards } \\ y r(n) \end{gathered}$ | $\begin{array}{r} 1 \mathrm{SW} \\ \operatorname{yr}(\mathrm{n}) \end{array}$ | Number in catch |  |  | . /rel |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | All SW | 2SW | Rec |  |
| 1984 | Faroe Islands | 19,620 | - | - | - | 9 | 9 | 0.46 |
|  | Ireland | 260,816 | 246 | - | 246 | 15 | 261 | 1.00 |
|  | N. Iceland | 72,352 | 33 | - | 33 | - | 33 | 0.45 |
|  | UK (Engl.+ Wales) | 39,780 | - | - | - | 3 | 3 | 0.08 |
|  | UK (Scotland) | 30,040 | 49 | - | 49 | - | 49 | 1.64 |
| 1985 | Faroe Islands | 30,079 | - | - | - | 87 | 87 | 2.89 |
|  | Ireland | 220,000 | 86 | - | 86 | 3 | 89 | 0.40 |
|  | UK (Engl.+ Wales) | 53,347 | - | - | - | 15 | 15 | 0.28 |
|  | UK (Scotland) | 13,497 | - | - | - | 3 | 3 | 0.22 |
| 1986 | Faroe Islands | 43,000 | - | - | - | 54 | 54 | 1.26 |
|  | Ireland | 143,866 | 30 | - | 30 | 11 | 41 | 0.29 |
|  | UK (Engl.+ Wales) | 177,071 | 4 | - | 4 | 8 | 12 | 0.07 |
|  | UK (N. Ireland) | 26,320 | 15 | - | 15 | - | 15 | 0.58 |
|  | UK (Scotland) | 16,217 | 8 | - | 8 | - | 8 | 0.47 |
| 1987 | Ireland | 162,189 | 154 | 3 | 157 | 4 | 161 | 0.99 |
|  | N. Iceland | 27,978 | - | 3 | 3 | 27 | 30 | 1.06 |
|  | UK (Engl.+ Wales) | 195,373 | 51 | - | 51 | 23 | 75 | 0.38 |
|  | UK (N. Ireland) | 20,145 | - | - | - | 2 | 2 | 0.09 |
|  | UK (Scotland) | 20,876 | - | - | - | 4 | 4 | 0.17 |
|  | USA | 640,400 | - | - | - | 2 | 2 | 0.00 |
| 1988 | Canada | 13,322 | 6 | - | - | - | 6 | 0.45 |
|  | Faroe Islands | 43,481 | 12 | - | 12 | 69 | 81 | 1.87 |
|  | Ireland | 165,841 | 104 | - | 104 | 7 | 111 | 0.67 |
|  | UK (Engl.+ Wales) | 189,913 | 12 | 2 | 14 | 12 | 26 | 0.13 |
|  | UK (Scotland) | 31,331 | 12 | - | 12 | 1 | 12 | 0.39 |
| 1989 | Faroe Islands | 26,943 | - | - | - | 8 | 8 | 0.28 |
|  | Ireland | 185,439 | 105 | - | 105 | - | 105 | 0.57 |
|  | N.Iceland | 85,452 | - | - | - | 4 | 45 | 0.04 |
|  | UK (Engl, \& Wales) | 256,342 | 23 | 2 | 25 | 15 | 40 | 0.16 |
|  | UK (Scotland) | 30,288 |  | 2 | 25 | 4 | 4 | 0.13 |
| 1990 | Ireland | 153,821 | 44 | - | 44 | NA | 44 | 0.28 |
|  | UK (Engl. \& Wales) | 250,024 | 15 | - | 15 | NA | 15 | 0.06 |
|  | UK (N. Ireland) | 29,875 | 15 | - | 15 | NA | 15 | 0.49 |
|  | UK (Scotland) | 41,390 | 15 | - | 15 | NA | 15 | 0.35 |

NA = not available

Table 4.1.6.1 Estimated exploitation rates of 1 SW and 2 SW salmon in the Faroes
fishery. Reporting rates for external tag recoveries assumed to be as follows: Faroese fishery 75\%, North Esk area 100\%, elsewhere in Scotland 75\%, Norwegian home water fisheries $50 \%$ and Sweden $65 \%$

Exploitation Rates \%

| Season | Norway |  |  |  |  |  | Scotland |  |  | Sweden |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | R. Drammen |  | R. Imsa |  |  |  | North Esk |  |  | R. Lagan |  |
|  | Hatchery |  | Wild |  | Hatchery |  | Wild |  |  | Hatchery |  |
|  | 1SW | 2SW | 1SW | 2SW | 1SW | 2SW | 1SW | 2SW | 3SW | 1SW | 2SW |
| 1981/1982 |  |  | 0 | - | 1 | - | 0 | 0 | 0 |  |  |
| 1982/1983 |  |  | 0 | 25 | 2 | 38 | <1 | 12 | 0 |  |  |
| 1983/1984 |  |  | 0 | 50 | 1 | 45 | 0 | 14 | 25 |  |  |
| 1984/1985 | 5 | - | 0 | 33 | 2 | 39 | 0 | 8 | 13 | 0 |  |
| 1985/1986 | 0 | 30 | 0 | 38 | 0 | 30 | <1 | 5 | 0 | 3 | 22 |
| 1986/1987 | 0 | 3 | 0 | 13 | 1 | 28 | 0 | 6 | 0 | 2 | 0 |
| 1987/1988 | 0 | 6 | 0 | 5 | 1 | 21 | 0 | 0 | 0 | 0 | 9 |
| 1988/1989 | 0 | 36 | 0 | 3 | 0 | 10 | 5 | 0 | 0 | 0 | 13 |
| 1989/1990 ${ }^{1}$ | 0 | 45 | 0 | 5 | 0 | 15 | 3 | 0 | 0 | 1 | 21 |
| 1990/1991 | 0 | 24 | 0 | 13 | 0 | 40 | <1 | 5 | 0 | 1 | 20 |

${ }^{1}$ Provisional exploitation rate estimates.

Table 4.2.4.1. Origin of catches of salmon in homewater fisheries.
$++=$ principal component of the catch
$+=$ other significant contributions

- = occurence

|  | Catch in Country |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Origin of Catch | Rus | Fin | Nor | Swe | Fr | UK E\&W | UK Scot | $\begin{array}{\|l\|} \hline \mathrm{UK} \\ \mathrm{NI} \\ \hline \end{array}$ | Ire | Ice |
| Russia | ++ | - | + |  |  |  |  |  |  |  |
| Finland |  | ++ | + |  |  |  |  |  |  |  |
| Norway |  | + | ++ | + |  | - | - |  | - |  |
| Sweden |  |  | + | ++ |  |  |  |  |  |  |
| France |  |  |  |  | ++ |  |  |  |  |  |
| UK (E\&W) |  |  | - | - | - | ++ | + | + | + |  |
| UK (Scot) |  |  |  |  |  | + | ++ | + | + |  |
| UK (NI) |  |  |  |  |  | - | + | ++ | + |  |
| Ireland |  |  | - | - | - | - | + | + | ++ |  |
| Iceland |  |  | - |  |  |  |  |  |  | ++ |

Table 4.2.4.2 Estimated catches (in tonnes round fresh weight) of wild, farmed and ranched salmon in homewater fisheries in 1991

| Country | Catches of salmon |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
|  | Wild | Farmed | Ranched | Total |
|  |  |  |  |  |
| Finland | 68 | $<1$ | 0 | 69 |
| France | $>12$ | 0 | $<1$ | 13 |
| Iceland | 122 | 3 | 394 | 519 |
| Ireland | $<422$ | + | 0 | 422 |
| Norway | 692 | 26 | (FW) | 0 |
|  |  | 167 | (Sea) | 885 |
| Russia | 215 | 0 | 0 | 215 |
| Sweden | 23 | 1 | $14^{1}$ | 38 |
| UK (E\&W) | 199 | 0 | 0 | 199 |
| UK (NI) | 54 | $<1$ | 0 | 55 |
| UK (Scot) | 384 | 12 | 0 | 396 |
|  |  |  |  |  |

${ }^{1}$ Fish released for mitigation and not expected to contribute to spawning stocks.

Table 4.2.5.1 Estimated exploitation rates (in \%) of salmon in homewater fisheries in Ireland and UK. Reporting rates for external tags shown below.

(1) Provisional figures.
(2) In-river netting only.

Reporting rates for external tags:

Table 4.2.5.2 Estimated exploitation rates (in \%) of salmon in homewater fisheries in Iceland, Norway, Russia, and Sweden. Reporting rates for external tags shown below.

|  | Iceland | Norway |  |  |  |  |  | Sweden |  | Russia |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & \text { R.Ellidaar } \\ & \text { wild } \end{aligned}$ |  | $\begin{aligned} & \text { ammen } \\ & \mathrm{HR}^{2} \end{aligned}$ |  | $\begin{aligned} & \text {.Ims } \\ & \text { ild } \end{aligned}$ | $\mathrm{HR}^{2}$ |  | $\begin{gathered} \text { Lagan }^{3} \\ \operatorname{HR}(2+) \end{gathered}$ |  | R. Ponoy R. KolaWild |  |
| Year | 1SW | 1SW | 2SW | ISW | 2SW | 1SW | 2SW | 1SW | 2SW | All sea ages | All sea |
| 1985 |  | 57 | - | 73 | 94 | 81 | 100 | 81 |  | 47 |  |
| 1986 | 34 | 81 | 50 | 79 | 82 | 78 | 90 | 93 | 82 | 50 | 91 |
| 1987 |  | 64 | 52 | 56 | 95 | 83 | 95 | 78 | 25 | 48 | 87 |
| 1988 |  | 70 | 47 | 51 | 80 | 78 | 91 | 73 | 82 | 77 | 84 |
| 1989 | 41 | 40 | 59 | 65 | 74 | 44 | 65 | 76 | 84 | 78 | 84 |
| 1990 | 44 | 5 | 40 | 60 | 42 | 52 | 68 | 80 | 82 | 50 | 80 |
| $1991{ }^{1}$ | 37 | 64 | 70 | 41 | 74 | 54 | 69 | 90 | 92 | 20 | 58 |
| Average | - 39 | 57 | 53 | 62 | 77 | 67 | 83 | 82 | 81 | 53 | 80 |

${ }^{1}$ Provisional figures.
${ }^{H R}$ in $R$. Drammen and $R$. Ims are pooled groups of $1+$ and $2+$ smolts.
${ }^{3}$ Assuming $50 \%$ exploitation in the river brood stock fishery

Reporting rates for external tags:
Faroes 75\%
Norwegian net fisheries $50 \%$
Sweden 65\%

Table 4.2.6.1 Nominal catches in Norwegian homewaters 1982-1991 ( $t$ round weight) broken down to drift net fishery, marine fishery excluding drift nets (other nets) and freshwater fishery and the proportion of the total catch taken in freshwater.

|  | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Drift nets | 590 | 826 | 866 | 667 | 795 | 552 | 527 | 0 | 0 | 0 |
| Other nets | 469 | 418 | 458 | 572 | 497 | 461 | 314 | 488 | 514 | 471 |
| Freshwater | 289 | 306 | 299 | 322 | 306 | 372 | 235 | 417 | 416 | 414 |
| Proportion in freshwater | 0.21 | 0.20 | 0.18 | 0.21 | 0.19 | 0.27 | 0.22 | 0.46 | 0.45 | 0.47 |
| Total | 1,348 | 1,550 | 1,623 | 1,561 | 1,598 | 1,385 | 1,076 | 905 | 930 | 885 |

* Provisional data

Table 4.2.6.2 Exploitation in Norwegian fisheries of 1 SW salmon from the River Lagan (Sweden). The estimates are based on $75 \%$ and $50 \%$ tag reporting rates in the Norwegian Sea and Norwegian homewaters, respectively.

| Year of fishery | 1 SW |
| :---: | :---: |
| 1985 | $5 \%$ |
| 1986 | $6 \%$ |
| 1987 | $5 \%$ |
| 1988 | $11 \%$ |
| 1989 | $0 \%$ |
| 1990 | $2 \%$ |
| 1991 | $1 \%$ |

Table 4.2.6.3 Frequency of net marks on Atlantic salmon in 10 Norwegian rivers sampled during 1978-1988 and in 1990-1991 (unweighted mean).

| River \% | 1978-1988 |  |  |  | 1990-1991 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number of sampling years | Total number of fish examined | $\begin{gathered} \text { Net marks } \\ \% \end{gathered}$ | $\begin{gathered} \text { Range } \\ \% \end{gathered}$ | Number of sampling years | Number of fish examined | $\begin{gathered} \text { Net marks } \\ \% \end{gathered}$ | Range |
| R. Rupperfiord | 7 | 4,812 | 29 | 18-45 | 1 | 165 | 29 | - |
| R. Malselv | 9 | 2,590 | 44 | 12-75 | 2 | 319 | 29 | 26-31 |
| R. Vefsna | 8 | 2,220 | 33 | 16-58 | 1 | 102 | 12 | - |
| R. Namsen | 9 | 4,036 | 25 | 12-36 | 2 | 367 | 10 | 4-15 |
| R. Stjordal | 4 | 889 | 43 | 32-63 | 2 | 237 | 15 | 6-24 |
| R. Orkla | 2 | 132 | 71 | 66-76 | 1 | 73 | 19 | - |
| R. Orsta | 7 | 2,094 | 73 | 48-90 | 1 | 138 | 20 | 17-23 |
| R. Gaular | 5 | 1,522 | 37 | 23-56 | 2 | 241 | 22 | 16-27 |
| R. Etne | 7 | 3,883 | 36 | 27-52 | 1 | 61 | 8 | - |
| R. Suldal | 7 | 1,025 | 18 | 8-43 | 2 | 505 | 2 | 1-3 |
| R. Imsa | 11 | 2,886 | 16 | 6-47 | 2 | 1,007 | 5 | 4-5 |
| R. Figgjo | 4 | 950 | 24 | 12-38 | 2 | 381 | 13 | 9-16 |

Table 5.1.1 Nominal catches at West Greenland, 1960-1991 (in tonnes, round fish weight).

| Year | Norway | Faroes | Sweden | Denmark | Greenland ${ }^{5}$ | Total | Quota |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | - | - | - | - | 60 | 60 | - |
| 1961 | - | - | - | - | 127 | 127 | - |
| 1962 | - | - | - | - | 244 | 244 | - |
| 1963 | - | - | - | - | 466 | 466 | - |
| 1964 | $\cdots$ | - | - | - | 1,539 | 1,539 | - |
| 1965 | - 1 | 36 | - | - | 825 | 861 | - |
| 1966 | 32 | 87 | - | - | 1,251 | 1,370 | - |
| 1967 | 78 | 155 | - | 85 | 1,283 | 1,601 | - |
| 1968 | 138 | 134 | 4 | 272 | 579 | 1,127 | - |
| 1969 | 250 | 215 | 30 | 355 | 1,360 | 2,210 | - |
| 1970 | 270 | 259 | 8 | 358 | 1,244 | 2,146 ${ }^{3}$ | - |
| 1971 | 340 | 255 | - | 645 | 1,449 | 2,689 | - |
| 1972 | 158 | 144 | - | 401 | 1,410 | 2,113 | - |
| 1973 | 200 | 171 | - | 385 | 1,585 | 2,341 | - |
| 1974 | 140 | 110 | - | 505 | 1,162 | 1,917 | - |
| 1975 | 217 | 260 | - | 382 | 1,171 | 2,030 | - |
| 1976 | - | - | - | - | 1,175 | 1,175 | 1,190 |
| 1977 | - | - | - | - | 1,420 | 1,420 | 1,190 |
| 1978 | - | - | - | - | 984 | 984 | 1,190 |
| 1979 | - | - | - | - | 1,395 | 1,395 | 1,190 |
| 1980 | - | - | - | - | 1,194 | 1,194 | 1,190 |
| 1981 | - | - | - | - | 1,264 | 1,264 | 1,265 ${ }^{5}$ |
| 1982 | - | - | - | - | 1,077 | 1,077 | 1,253 ${ }^{5}$ |
| 1983 | - | - | - | - | 310 | 310 | 1,190 |
| 1984 | - | - | - | - | 297 | 297 | 870 |
| 1985 | - | - | - | - | 864 | 864 | 852 |
| 1986 | - | - | - | - | 960 | 960 | 909 |
| 1987 | - | - | - | - | 966 | 966 | 935 |
| 1988 | - | - | - | - | 893 | 893 | - |
| 1989 | - | - | - | - | 337 | 337 | 900 |
| 1990 | - | - | - | - | 227 | 227 | 924 |
| 1991 | - | - | - - | - | 437 | $437{ }^{2}$ | 840 |

Figures not available, but catch is known to be less than the Faroese catch.
${ }_{2}^{2}$ Provisional.
${ }^{3}$ Including 7 t caught on longline by one of two Greenland vessels in the Labrador Sea early in 1970.
${ }^{4}$ For Greenlandic vessels: all catches up to 1968 were taken with set gillnets only; after 1968, the catches were taken with set gillnets and drift nets. All non-Greenlandic catches from 1969-1984 were taken with drift nets. ${ }^{5}$ Quota corresponding to specific opening dates of the fishery.

Factor used for converting landed catch to round fresh weight in fishery by Greenland vessels $=1.11$. Factor for Norwegian, Danish, and Faroese drift net vessels $=1.10$.

Table 5.1.2 Distribution of nominal catches (tonnes) taken by Greenland vessels in 1977-1991 by NAFO divisions according to place where landed.
Div. $\begin{array}{lllllllllllllllllll}1977 & 1978 & 1979 & 1980 & 1981 & 1982 & 1983 & 1984 & 1985 & 1986 & 1987 & 1988 & 1989 & 1990 & 1991\end{array}$

| 1A | 201 | 81 | 120 | 52 | 105 | 111 | 14 | 33 | 85 | 46 | 48 | 24 | 9 | 3 | 22 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1B | 393 | 349 | 343 | 275 | 403 | 330 | 77 | 116 | 124 | 73 | 114 | 100 | 28 | 18 | 21 |
| 1 C | 336 | 245 | 524 | 404 | 348 | 239 | 93 | 64 | 198 | 128 | 229 | 213 | 81 | 99 | 109 |
| 1D | 207 | 186 | 213 | 231 | 203 | 136 | 41 | 4 | 207 | 203 | 205 | 191 | 73 | 44 | 31 |
| 1E | 237 | 113 | 164 | 158 | 153 | 167 | 55 | 43 | 147 | 233 | 261 | 198 | 75 | 34 | 106 |
| 1F | 46 | 10 | 31 | 74 | 32 | 76 | 30 | 32 | 103 | 277 | 109 | 167 | 71 | 29 | 148 |
| 1NK | - | - | - | - | 20 | 18 | - | 5 | - | - | - | - | - | - | - |
| Total | 1,420 | 984 | 1,395 | 1,194 | 1,264 | 1,077 | 310 | 297 | 864 | 960 | 966 | 893 | 337 | 227 | 437 |

East

$\begin{array}{lllllllllllllllllllll}\text { Total } 1,426 & 992 & 1,395 & 1,194 & 1,264 & 1,077 & 310 & 297 & 871 & 979 & 966 & 897 & 337 & 227 & 441\end{array}$

[^3]Table 5.1.1.1. Percentage (by number) of North American and European salmon in research vessel catches at West Greenland (1969-1982) and from commercial samples (1978-1991).

| Source | Year | Sample size |  | Continent of origin (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Length | Scales | NA | $(95 \% \mathrm{CI})^{1}$ | E | (95\% CI) |
| Research | 1969 | 212 | 212 | 51 | $(57,44)$ | 49 | $(56,43)$ |
|  | 1970 | 127 | 127 | 35 | $(43,26)$ | 65 | $(74,57)$ |
|  | 1971 | 247 | 247 | 34 | $(40,28)$ | 66 | $(72,50)$ |
|  | 1972 | 3,488 | 3,488 | 36 | $(37,34)$ | 64 | $(66,63)$ |
|  | 1973 | 102 | 102 | 49 | $(59,39)$ | 51 | $(61,41)$ |
|  | 1974 | 834 | 834 | 43 | $(46,39)$ | 57 | $(61,54)$ |
|  | 1975 | 528 | 528 | 44 | $(48,40)$ | 56 | $(60,52)$ |
|  | 1976 | 420 | 420 | 43 | $(48,38)$ | 57 | $(62,52)$ |
|  | 1977 | - | - | - | ( - ) | - | ( - ) |
|  | $1978{ }^{2}$ | 606 | 606 | 39 | $(41,34)$ | 62 | $(66,59)$ |
|  | $1978{ }^{3}$ | 49 | 49 | 55 | $(69,41)$ | 45 | $(59,31)$ |
|  | 1979 | 328 | 328 | 47 | $(52,41)$ | 53 | $(59,48)$ |
|  | 1980 | 617 | 617 | 58 | $(62,54)$ | 42 | $(46,38)$ |
|  | 1981 | - | - | - | $(-)$ | - | ( - ) |
|  | 1982 | 443 | 443 | 47 | $(52,43)$ | 53 | $(58,48)$ |
| Commercial | 1978 | 392 | 392 | 52 | $(57,47)$ | 48 | $(53,43)$ |
|  | 1979 | 1,653 | 1,653 | 50 | $(52,48)$ | 50 | $(52,48)$ |
|  | 1980 | 978 | 978 | 48 | $(51,45)$ | 52 | $(55,49)$ |
|  | 1981 | 4,570 | 1,930 | 59 | $(61,58)$ | 41 | $(42,39)$ |
|  | 1982 | 1,949 | 414 | 62 | $(64,60)$ | 38 | $(40,36)$ |
|  | 1983 | 4,896 | 1,815 | 40 | $(41,38)$ | 60 | $(62,59)$ |
|  | 1984 | 7,282 | 2,720 | 50 | $(53,47)$ | 50 | $(53,47)$ |
|  | 1985 | 13,272 | 2,917 | 50 | $(53,46)$ | 50 | $(54,47)$ |
|  | 1986 | 20,394 | 3,509 | 57 | $(66,48)$ | 43 | $(52,34)$ |
|  | 1987 | 13,425 | 2,960 | 59 | $(63,54)$ | 41 | $(46,37)$ |
|  | 1988 | 11,047 | 2,562 | 43 | $(49,38)$ | 57 | $(62,51)$ |
|  | 1989 | 9,366 | 2,227 | 56 | $(60,52)$ | 44 | $(48,40)$ |
|  | 1990 | 4,897 | 1,208 | 75 | $(79,70)$ | 25 | $(30,21)$ |
|  | 1991 | 5,011 | 1,347 | 65 | $(69,61)$ | 35 | $(39,31)$ |

${ }^{1}$ CI - Confidence Interval calculated by method of Pella and Robertson (1989) for 1984-1986 and by binominal distribution for the others)
${ }^{2}$ During fishery.
${ }^{3}$ Research samples after fishery closed.

Table 5.1.1.2. The numbers of North American and European Atlantic salmon caught at West Greenland from 1982 to 1991. Number are rounded to the nearest hundred fish.

|  | Numbers of |  |
| :--- | :---: | :---: |
| Year | NA | Salmon Caught |
| 1982 | 192,200 | 143,800 |
| 1983 | 39,500 | 60,500 |
| 1984 | 48,800 | 42,200 |
| 1985 | 143,500 | 161,500 |
| 1986 | 188,300 | 131,900 |
| 1987 | 171,900 | 126,400 |
| 1988 | 125,500 | 168,800 |
| 1989 | 65,000 | 52,700 |
| 1990 | 62,400 | 21,700 |
| 1991 | 103,000 | 60,900 |

Table 5.1.1.3 The number and percentage of fin-clipped and microtagged Atlantic salmon observed during the sampling programme at West Greenland in 1991.

| SAMPLING <br> SITE <br> (sampling <br> period) | NAFO DIV. | NO. SALMON EXAMINED |  |  |  | NO. AFC's OBSERVED |  |  |  | \% AFC'S | NO.CWT's RECOVERED |  |  |  | \% CWT's | $\begin{gathered} \% \text { AFC's } \\ \text { with } \\ \text { CWT's } \end{gathered}$ | No.untagged AFC Fish | untagged |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $<3 \mathrm{Kg} \mathrm{3-5} \mathrm{Kg}>5 \mathrm{Kg}$ Total |  |  |  | $<3 \mathrm{Kg} 3-5 \mathrm{Kg}>5 \mathrm{Kg}$ Total |  |  |  |  | $\begin{aligned} & \mathrm{V} \\ & <3 \mathrm{Kg} \end{aligned}$ | $\begin{aligned} & \text { Veight } \\ & 3-5 \mathrm{Kg} \end{aligned}$ |  | Total |  |  |  |  |
| MANIITSOQ <br> (10-28 Aug) | 1 C | 3688 |  | 52 | 3871 |  | 5 | 0 | 72 | 1.86 |  | 1 | 0 | 18 | 0.46 | 25.0 | 54 | 1.39 |
| NUUK (14-23 Aug) | 1D | 2232 | 64 | 28 | 2324 |  |  | 0 | 30 | 1.29 |  | 0 | 0 | 12 | 0.52 | 40.0 | 18 | 0.77 |
| $\begin{aligned} & \text { PAAMIUT } \\ & (13-25 \text { Aug }) \end{aligned}$ | 1E | 983 | 34 | 17 | 1162 | 16 | 1 | 0 | 19 | 1.64 | 6 | 1 | 0 | 7 | 0.60 | 36.8 | 12 | 1.03 |
| TOTAL |  | 6903 | 229 | 97 | 7357 |  |  | 0 | 121 | 1.64 |  |  | 0 | 37 | 0.50 | 30.6 | 84 | 1.14 |
| \% |  | 95.49 | 3.17 | 1.34 |  | 94.12 | 5.88 | 0 |  |  | 94.59 | 5.41 | 0 |  |  |  |  |  |

```
AFC = Adipose fin clip.
CWT = Coded wire microtag
```

Table 5.1.1.4 Distributions of recaptures and recovery rates (per 1000 fish examined) of microtagged salmon from Europe and N.America caught in the West Greenland fishery, 1986-1991

| SCANNING |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | No. Scanned by NaFO Division |  |  |  |  |  |
|  | 1A | 1B | 1 C | 1D | 1E | 1 F |
| 1986 | - | 7308 | - | 10120 | 7361 | 5571 |
| 1987 | - | 6073 | - | 9263 | 6139 | 3572 |
| 1988 | 522 | 7410 | - | 9884 | 2523 | 2258 |
| 1989 | - | 3654 | - | 7591 | 4343 | - |
| 1990 | - | - | 1156 | 3827 | 1427 | - |
| 1991 | - | - | 3871 | 2324 | 1162 | - |
| N. AMERICAN RECOVERIES |  |  |  |  |  |  |
| TOT. | No. Recovs by NaFO Division |  |  |  |  |  |
| RECS | 1A | 1B | 1 C | 1D | 1E | 1 F |
| 198626 | - | 10 | - | 11 | 4 | 1 |
| 1987103 | - | 33 | - | 43 | 11 | 16 |
| 198881 | 2 | 25 | - | 40 | 12 | 2 |
| 198972 | - | 31 | - | 34 | 7 | - |
| 199046 | - | - | 16 | 29 | 1 | - |
| 199128 | - | - | 14 | 9 | 5 | - |
|  | Recovery rate (per 1000 fish) |  |  |  |  |  |
|  | 1A | 1B | 1 C | 1D | 1 E | 1F |
| 1986 | - | 1.4 | - | 1.1 | 0.5 | 0.2 |
| 1987 | - | 5.4 | - | 4.6 | 1.8 | 4.5 |
| 1988 | 3.8 | 3.4 | - | 4.0 | 4.8 | 0.9 |
| 1989 | - | 8.5 | - | 4.5 | 1.6 | - |
| 1990 | - | - | 13.8 | 7.6 | 0.7 | - |
| 1991 | - | - | 3.6 | 3.9 | 4.3 | - |
| EUROPEAN RECOVERIES |  |  |  |  |  |  |
| TOT. | No. | Recovs by NAFO Division |  |  |  |  |
| RECS |  | 1B | 1 C | 1D | 1E | 1 F |
| 198622 | - | 9 | - | 10 | 3 | 0 |
| 198743 | - | 13 | - | 18 | 7 | 5 |
| 198829 | 1 | 10 | - | 11 | 6 | 1 |
| 198927 | - | 10 | - | 10 | 7 | - |
| 1990 8 | - | - | 1 | 3 | 4 | - |
| 19916 | - | - | 3 | 1 | 2 | - |
| Recovery rate (per 1000 fish) |  |  |  |  |  |  |
|  | 1A | 1B | 1 C | 1D | 1E | 1F |
| 1986 | - | 1.2 | - | 1.0 | 0.4 | 0.0 |
| 1987 | - | 2.1 | - | 1.9 | 1.1 | 1.4 |
| 1988 | 1.9 | 1.3 | - | 1.1 | 2.4 | 0.4 |
| 1989 | - | 2.7 | - | 1.3 | 1.6 | - |
| 1990 | - | - | 0.9 | 0.8 | 2.8 | - |
| 1991 | - | - | 0.8 | 0.4 | 1.7 | - |
| A dash represents no scanning undertaken. |  |  |  |  |  |  |

Table 5.1.1.5 Mean lengths, weights and distributions of recaptures at West Greenland in1991 of microtagged salmon from different release areas. Recovery rates per 1000 fish examined are also given in parenthesis for each NAFO division.

| COUNTRY | $\begin{gathered} \text { RELEASE } \\ \text { SITE } \end{gathered}$ | NUMBER RELEASED | RELEASE STAGE | TOT. NO. RECOVS | RECOVS. BY NAFO DIV. <br> 1C 1 I |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| USA | Connecticut R <br> Penobscot R. <br> Merrimack R. <br> Total <br> Recovery rate | 875,306 | $\begin{aligned} & \mathrm{HS} \\ & \mathrm{HS} \\ & \mathrm{HS} \\ & \mathrm{HS} \end{aligned}$ | $\begin{gathered} 4 \\ 19 \\ 3 \\ 26 \\ 3.5 \end{gathered}$ | $\begin{gathered} 2 \\ 10 \\ 2 \\ 14 \\ 3.6 \end{gathered}$ | $\begin{gathered} 2 \\ 6 \\ 0 \\ 8 \\ 3.4 \end{gathered}$ | $\begin{gathered} 0 \\ 3 \\ 1 \\ 4 \\ 3.4 \end{gathered}$ |
| CANADA | Nepisiguit R. Stewart Brook, N.B <br> Total <br> Recovery rate | $\begin{aligned} & 74,329 \\ & 30,384 \end{aligned}$ | $\begin{aligned} & \text { HS } \\ & \text { HS } \\ & \text { HS } \\ & \text { WS } \end{aligned}$ | $\begin{gathered} 1 \\ 1 \\ 2 \\ 0.3 \end{gathered}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{gathered} 1 \\ 0 \\ 1 \\ 0.4 \end{gathered}$ | $\begin{gathered} 0 \\ 1 \\ 1 \\ 0.9 \end{gathered}$ |
| IRELAND | Castleconnell R. <br> R. Lee <br> Total <br> Recovery rate | $\begin{gathered} 151,469 \\ 16,789 \end{gathered}$ | HS <br> HS <br> HS <br> WS | $\begin{gathered} 1 \\ 1 \\ 2 \\ 0.3 \end{gathered}$ | $\begin{gathered} 1 \\ 0 \\ 1 \\ 0.3 \end{gathered}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{gathered} 0 \\ 1 \\ 1 \\ 0.9 \end{gathered}$ |
| ENGLAND \& WALES | R. Wear <br> R. Taff <br> Total <br> Recovery rate | $\begin{gathered} 241,361 \\ 8,663 \end{gathered}$ | $\begin{gathered} \text { WS } \\ \text { HS } \\ H S / H P \\ \text { WS } \end{gathered}$ | $\begin{gathered} 2 \\ 1 \\ 3 \\ 0.4 \end{gathered}$ | $\begin{gathered} 1 \\ 0 \\ 1 \\ 0.3 \end{gathered}$ | $\begin{gathered} 1 \\ 1 \\ 2 \\ 0.9 \end{gathered}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ |
| ICELAND | Hraunsfjordur Bruara Total <br> Recovery rate | $\begin{gathered} 405,305 \\ 6,000 \end{gathered}$ | $\begin{aligned} & \text { HS } \\ & \text { HS } \\ & \text { HS } \\ & \text { WS } \end{aligned}$ | $\begin{gathered} 1 \\ 1 \\ 2 \\ 0.3 \end{gathered}$ | $\begin{gathered} 0 \\ 1 \\ 1 \\ 0.3 \end{gathered}$ | $\begin{gathered} 1 \\ 0 \\ 1 \\ 0.4 \end{gathered}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ |
| SCOTLAND | Girnock Burn <br> Recovery rate | $\begin{gathered} 33,460 \\ 7,930 \end{gathered}$ | $\begin{aligned} & \text { HS } \\ & \text { WS } \end{aligned}$ | $\begin{gathered} 1 \\ 0.1 \end{gathered}$ | 0 | 0 0 | 1 0.9 |
| N.IRELAND | R.Bush <br> Recovery rate | $\begin{gathered} 27,076 \\ 2,799 \end{gathered}$ | $\begin{aligned} & \text { HP } \\ & \text { WS } \end{aligned}$ | $\begin{gathered} 1 \\ 0.1 \end{gathered}$ | 1 0.3 | 0 0 | 0 0 |

KEY: $\quad$ ' 90 HS $=1990$ Hatchery smolt release.
'90 WS = 1990 Wild smolt release .
'90 HP = 1990 Hatchery parr release.

Table 5.1.1.6 Summary of input data for harvest calculations: Releases by smolt age for Maine and Canada.

| Re- <br> lease <br> Year | Maine Rivers |  | $\qquad$ <br> Total: 1-YR | Snolt <br> Multiplier <br> f1*f2 | $\xrightarrow[\text { Canada }]{\substack{\text { C- } \\ \hline \text { Yr (Cl) }}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |
|  | 1-YR (U1) | 2-YR (U2) |  |  |  |
| 1975 | 15758 | 153577 | 10.75 | 3.809 | 28700 |
| 1976 | 60229 | 242468 | 5.03 | 1.980 | 92636 |
| 1977 | 128885 | 245608 | 2.91 | 1.403 | 138000 |
| 1978 | 168033 | 135014 | 1.80 | 1.007 | 132900 |
| 1979 | 98693 | 272585 | 3.76 | 2.342 | 59800 |
| 1980 | 399903 | 282001 | 1.71 | 1.295 | 126300 |
| 1981 | 24695 | 232348 | 10.41 | 2.098 | 97800 |
| 1982 | 135007 | 259674 | 2.92 | 1.525 | 123700 |
| 1983 | 367605 | 170277 | 1.46 | 0.916 | 219200 |
| 1984 | 657722 | 137203 | 1.21 | 0.871 | 254800 |
| 1985 | 612548 | 108598 | 1.18 | 0.838 | 247400 |
| 1986 | 723400 | 55000 | 1.08 | 0.661 | 452800 |
| 1987 | 637536 | 82759 | 1.13 | 0.662 | 449300 |
| 1988 | 850900 | 87100 | 1.10 | 0.708 | 472500 |
| 1989 | 524300 | 80200 | 1.15 | 0.571 | 533954 |
| 1990 | - 644100 | 33100 | 1.05 | 0.537 | 617353 |

"Canadian smolt releases adjusted to include estimated "smolts" that contribute to West Greenland.

Table 5.1.1.7 Summary of input data for harvest calculations: Mean weights, landings, proportion North American stock, and the fraction of river age 1 harvest in Greenland.

| YEAR |  | Mean <br> Weight | Std Dev Weight | Prop. <br> NA | Sample | Number of <br> NA 1-yr | Harvest | Estimates - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{CATCH}$ |  |  |  | Size |  | Grld | North American |
|  | mt | (kg) | (kg) | Origin | for Prop. | in sample | no. | Total |
| i | Q | IT | SD(W) | $p_{\text {na }}$ | n | $\mathrm{n}_{1}$ | C | $\mathrm{p}_{\mathrm{na}} \mathrm{C}$ |
| 1976 | 1175 | 3.04 | 0.98 | 0.44 | 275 | 1 | 386513 | 170066 |
| 1977 | 1420 | 3.21 | 0.96 | 0.43 | 576 | 2 | 442023 | 190070 |
| 1978 | 984 | 3.35 | 0.88 | 0.45 | 449 | 12 | 293731 | 132546 |
| 1979 | 1395 | 3.34 | 1.19 | 0.45 | 1358 | 57 | 417665 | 187949 |
| 1980 | 1194 | 3.22 | 0.88 | 0.53 | 1111 | 65 | 370807 | 194674 |
| 1981 | 1264 | 3.17 | 0.89 | 0.48 | 1109 | 39 | 398738 | 189401 |
| 1982 | 1077 | 3.11 | 0.96 | 0.57 | 446 | 6 | 346302 | 197392 |
| 1983 | 310 | 3.10 | 0.98 | 0.40 | 779 | 24 | 100000 | 40000 |
| 1984 | 297 | 3.11 | 1.14 | 0.54 | 1368 | 66 | 95498 | 51569 |
| 1985 | 864 | 2.87 | 0.86 | 0.47 | 1410 | 72 | 301045 | 141491 |
| 1986 | 960 | 3.03 | 0.85 | 0.59 | 1786 | 36 | 316832 | 186931 |
| 1987 | 966 | 3.16 | 0.74 | 0.59 | 1714 | 69 | 305696 | 180361 |
| 1988 | 893 | 3.18 | 0.77 | 0.40 | 1025 | 53 | 280818 | 112327 |
| 1989 | 337 | 2.87 | 0.89 | 0.56 | 1025 | 80 | 117422 | 65756 |
| 1990 | 227 | 2.69 | 0.83 | 0.75 | 1025 | 90 | 84387 | 63290 |
| 1991 | 440 | 2.65 | 0.92 | 0.65 | 771 | 40 | 166038 | 107925 |

Table 5.1.1.8. Harvest estimates in numbers of Maine salnon in West Greenland using the proportional harves nodel. Estimates include a non-catch fishing mortality rate of 0.2 .

| Year | Total <br> NA $1-\mathrm{Yr}$ | Total Maine RA $1+2$ | Std Err | Mean-2SE | Mean+2SE | 100 SE/Mean |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1976 | 1,237 | 5,889 | 3,130 | -371 | 12,149 | 53.15 |
| 1977 | 3,820 | 9,456 | 1,345 | 6,766 | 12,146 | 14.22 |
| 1978 | 3,542 | 6,213 | 861 | 4,492 | 7,935 | 13.85 |
| 1979 | 7,889 | 9,930 | 430 | 9,070 | 10,790 | 4.33 |
| 1980 | 11,390 | 33,351 | 1,305 | 30,741 | 35,961 | 3.91 |
| 1981 | 6,661 | 10,789 | 618 | 9,553 | 12,026 | 5.73 |
| 1982 | 2,656 | 6,965 | 1,623 | 3,719 | 10,212 | 23.31 |
| 1983 | 1,232 | 2,350 | 194 | 1,962 | 2,738 | 8.25 |
| 1984 | 2,488 | 2,851 | 111 | 2,629 | 3,073 | 3.89 |
| 1985 | 7,225 | 7,867 | 286 | 7,296 | 8,439 | 3.63 |
| 1986 | 3,768 | 3,950 | 240 | 3,469 | 4,430 | 6.08 |
| 1987 | 7,261 | 6,006 | 224 | 5,558 | 6,455 | 3.73 |
| 1988 | 5,808 | 4,812 | 219 | 4,374 | 5,249 | 4.55 |
| 1989 | 5,132 | 4,547 | 153 | 4,241 | 4,853 | 3.37 |
| 1990 | 5,557 | 3,968 | 123 | 3,722 | 4,214 | 3.10 |
| 1991 | 5,599 | 3,757 | 212 | 3,333 | 4,182 | 5.65 |

Table 5.1.2.1 Annual mean fork lengths and whole weights of Atlantic salmon caught at West Greenland, 1969-1991. Fork length (cm); whole weight (kg). NA = North American; E = European.

|  | Whole weight (kg) |  |  |  |  |  |  |  |  | Fork length (cm) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Sea age |  |  |  |  |  |  |  |  | Sea age |  |  |  |  |  |
|  | 1SW |  | 2SW |  | PS |  | Total |  | Total | 1SW |  | 2SW |  | PS |  |
|  | NA | E | NA | E | NA | E | NA | E |  | NA | E | NA | E | NA | E |
| 1969 | 3.12 | 3.76 | 5.48 | 5.80 | - | 5.13 | 3.25 | 3.86 | 3.58 | 65.0 | 68.7 | 77.0 | 80.3 |  | 75.3 |
| 1970 | 2.85 | 3.46 | 5.65 | 5.50 | 4.85 | 3.80 | 3.06 | 3.53 | 3.28 | 64.7 | 68.6 | 81.5 | 82.0 | 78.0 | 75.0 |
| 1971 | 2.65 | 3.38 | 4.30 | - | - | - | 2.68 | 3.38 | 3.14 | 62.8 | 67.7 | 72.0 | - | - |  |
| 1972 | 2.96 | 3.46 | 5.85 | 6.13 | 2.65 | 4.00 | 3.25 | 3.55 | 3.44 | 64.2 | 67.9 | 80.7 | 82.4 | 61.5 | 69.0 |
| 1973 | 3.28 | 4.54 | 9.47 | 10.00 | - | - | 3.83 | 4.66 | 4.18 | 64.5 | 70.4 | 88.0 | 96.0 | 61.5 |  |
| 1974 | 3.12 | 3.81 | 7.06 | 8.06 | 3.42 | - | 3.22 | 3.86 | 3.58 | 64.1 | 68.1 | 82.8 | 87.4 | 66.0 |  |
| 1975 | 2.58 | 3.42 | 6.12 | 6.23 | 2.60 | 4.80 | 2.65 | 3.48 | 3.12 | 61.7 | 67.5 | 80.6 | 82.2 | 66.0 | 75.0 |
| 1976 | 2.55 | 3.21 | 6.16 | 7.20 | 3.55 | 3.57 | 2.75 | 3.24 | 3.04 | 61.3 | 65.9 | 80.7 | 87.5 | 72.0 | 70.7 |
| 1977 | - | - | - | - | - | - | - | - | - |  |  |  | - | - |  |
| 1978 | 2.96 | 3.50 | 7.00 | 7.90 | 2.45 | 6.60 | 3.04 | 3.53 | 3.35 | 63.7 | 67.3 | 83.6 | - | 60.8 | 85.0 |
| 1979 | 2.98 | 3.50 | 7.06 | 7.60 | 3.92 | 6.33 | 3.12 | 3.56 | 3.34 | 63.4 | 66.7 | 81.6 | 85.3 | 61.9 | 82.0 |
| 1980 | 2.98 | 3.33 | 6.82 | 6.73 | 3.55 | 3.90 | 3.07 | 3.38 | 3.22 | 64.0 | 66.3 | 82.9 | 83.0 | 67.0 | 70.9 |
| 1981 | 2.77 | 3.48 | 6.93 | 7.42 | 4.12 | 3.65 | 2.89 | 3.58 | 3.17 | 62.3 | 66.7 | 82.8 | 84.5 | 72.5 | - |
| 1982 | 2.79 | 3.21 | 5.59 | 5.59 | 3.96 | 5.66 | 2.92 | 3.43 | 3.11 | 62.7 | 66.2 | 78.4 | 77.8 | 71.4 | 80.9 |
| 1983 | 2.54 | 3.01 | 5.79 | 5.86 | 3.37 | 3.55 | 3.02 | 3.14 | 3.10 | 61.5 | 65.4 | 81.1 | 81.5 | 68.2 | 70.5 |
| 1984 | 2.64 | 2.84 | 5.84 | 5.77 | 3.62 | 5.78 | 3.20 | 3.03 | 3.11 | 62.3 | 63.9 | 80.7 | 80.0 | 69.8 | 79.5 |
| 1985 | 2.50 | 2.89 | 5.42 | 5.45 | 5.20 | 4.97 | 2.72 | 3.01 | 2.87 | 61.2 | 64.3 | 78.9 | 78.6 | 79.1 | 77.0 |
| 1986 | 2.75 | 3.13 | 6.44 | 6.08 | 3.32 | 4.37 | 2.89 | 3.19 | 3.03 | 62.8 | 65.1 | 80.7 | 79.8 | 66.5 | 73.4 |
| 1987 | 3.00 | 3.20 | 6.36 | 5.96 | 4.69 | 4.70 | 3.10 | 3.26 | 3.16 | 64.2 | 65.6 | 81.2 | 79.6 | 74.8 | 74.8 |
| 1988 | 2.83 | 3.36 | 6.77 | 6.78 | 4.75 | 4.64 | 2.93 | 3.41 | 3.18 | 63.0 | 66.6 | 82.1 | 82.4 | 74.7 | 73.8 |
| 1989 | 2.56 | 2.86 | 5.87 | 5.77 | 4.23 | 5.83 | 2.77 | 2.99 | 2.87 | 62.3 | 64.5 | 80.8 | 81.0 | 73.8 | 82.2 |
| 1990 | 2.53 | 2.61 | 6.47 | 5.78 | 3.90 | 5.09 | 2.67 | 2.72 | 2.64 | 62.3 | 62.7 | 83.4 | 81.1 | 72.6 | 78.6 |
| 1991 | 2.42 | 2.54 | 5.82 | 6.23 | 5.15 | 5.09 | 2.57 | 2.79 | 2.65 | 61.6 | 62.7 | 80.6 | 82.2 |  | 80.0 |

Table 5.1.2.2 Sea age composition (\%) from research vessel and commercial catch samples of Atlantic salmon at West Greenland, 1969-1991.

| Year | Type | 1SW | MSW | PS |
| :--- | :--- | :--- | :---: | :---: |
|  |  |  |  |  |
| 1969 | Research | 93.8 | 4.9 | 1.3 |
| 1970 | Research | 93.8 | 4.1 | 2.1 |
| 1971 | Research | 99.2 | 0.4 | 0.4 |
| 1972 | Research | 94.1 | 5.6 | 0.3 |
| 1973 | Research | 93.8 | 4.4 | 1.8 |
| 1974 | Research | 97.7 | 1.7 | 0.6 |
| 1975 | Research | 97.6 | 2.0 | 0.4 |
| 1976 | Research | 95.7 | 2.6 | 1.7 |
| 1977 | No observations | 96.9 |  | - |
| 1978 | Research | 96.6 | 1.1 | 1.1 |
| 1979 | Commercial | 96.7 | 2.1 | 1.3 |
|  | Research | 97.5 | 1.8 | 1.5 |
| 1980 | Commercial | 98.4 | 2.2 | 0.3 |
|  | Research | 97.0 | 1.1 | 0.5 |
| 1981 | Commercial | 93.6 | 2.5 | 0.6 |
| 1982 | Commercial | 95.3 | 6.0 | 0.5 |
|  | Research | 90.5 | 2.4 | 2.2 |
| 1983 | Commercial | 87.6 | 8.1 | 1.4 |
| 1984 | Commercial | 93.8 | 5.6 | 0.7 |
| 1985 | Commercial | 96.2 | 3.9 | 0.3 |
| 1986 | Commercial | 97.0 | 2.0 | 0.8 |
| 1987 | Commercial | 97.4 | 1.7 | 1.0 |
| 1988 | Commercial | 93.8 | 4.6 | 1.9 |
| 1989 | Commercial | 95.9 | 3.2 | 0.9 |
| 1990 | Commercial | 94.7 | 4.9 | 0.3 |
| 1991 | Commercial |  |  |  |

Table 5.1.2.3 The sea age composition of samples from commercial catches at West Greenland, 1985-1991.

| Year | Sea age composition (\%) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North American |  |  | European |  |  |
|  | 1SW | 2SW | Previous spawners | 1SW | 2SW | Previous spawners |
| 1985 | 92.5 | 7.2 | 0.3 | 95.0 | 4.7 | 0.4 |
| 1986 | 95.1 | 3.9 | 1.0 | 97.5 | 1.9 | 0.6 |
| 1987 | 96.3 | 2.3 | 1.4 | 98.0 | 1.7 | 0.3 |
| 1988 | 96.7 | 2.0 | 1.2 | 98.1 | 1.3 | 0.5 |
| 1989 | 92.3 | 5.2 | 2.4 | 95.5 | 3.8 | 0.6 |
| 1990 | 95.7 | 3.4 | 0.9 | 96.3 | 3.0 | 0.7 |
| 1991 | 95.6 | 4.1 | 0.4 | 93.4 | 6.4 | 0.2 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| North American |  |  |  |  |  |  |  |  |
| 1968 | 0.3 | 19.6 | 40.4 | 21.3 | 16.2 | 2.2 | 0.0 | 0.0 |
| 1969 | 0.0 | 27.1 | 45.8 | 19.6 | 6.5 | 0.9 | 0.0 | 0.0 |
| 1970 | 0.0 | 58.1 | 25.6 | 11.6 | 2.3 | 2.3 | 0.0 | 0.0 |
| 1971 | 1.2 | 32.9 | 36.5 | 16.5 | 9.4 | 3.5 | 0.0 | 0.0 |
| 1972 | 0.8 | 31.9 | 51.4 | 10.6 | 3.9 | 1.2 | 4.0 | 0.0 |
| 1973 | 2.0 | 40.8 | 34.7 | 18.4 | 2.0 | 2.0 | 0.0 | 0.0 |
| 1974 | 0.9 | 36.0 | 36.6 | 12.0 | 11.7 | 2.6 | 0.3 | 0.0 |
| 1975 | 0.4 | 17.3 | 47.6 | 24.4 | 6.2 | 4.0 | 0.0 | 0.0 |
| 1976 | 0.7 | 42.6 | 30.6 | 14.6 | 10.9 | 0.4 | 0.4 | 0.0 |
| 1977 | - | - | - | - |  |  | - | - |
| 1978 | 2.7 | 31.9 | 43.0 | 13.6 | 6.0 | 2.0 | 0.9 | 0.0 |
| 1979 | 4.2 | 39.9 | 40.6 | 11.3 | 2.8 | 1.1 | 0.1 | 0.0 |
| 1980 | 5.9 | 36.3 | 32.9 | 16.3 | 7.9 | 0.7 | 0.1 | 0.0 |
| 1981 | 3.5 | 31.6 | 37.5 | 19.0 | 6.6 | 1.6 | 0.2 | 0.0 |
| 1982 | 1.4 | 37.7 | 38.3 | 15.9 | 5.8 | 0.7 | 0.0 | 0.2 |
| 1983 | 3.1 | 47.0 | 32.6 | 12.7 | 3.7 | 0.8 | 0.1 | 0.0 |
| 1984 | 4.8 | 51.7 | 28.9 | 9.0 | 4.6 | 0.9 | 0.2 | 0.0 |
| 1985 | 5.1 | 41.0 | 35.7 | 12.1 | 4.9 | 1.1 | 0.1 | 0.0 |
| 1986 | 2.0 | 39.9 | 33.4 | 20.0 | 4.0 | 0.7 | 0.0 | 0.0 |
| 1987 | 3.9 | 41.4 | 31.8 | 16.7 | 5.8 | 0.4 | 0.0 | 0.0 |
| 1988 | 5.2 | 31.3 | 30.8 | 20.9 | 10.7 | 1.0 | 0.1 | 0.0 |
| 1989 | 7.9 | 39.0 | 30.1 | 15.9 | 5.9 | 1.3 | 0.0 | 0.0 |
| 1990 | 8.8 | 45.3 | 30.7 | 12.1 | 2.4 | 0.5 | 0.1 | 0.0 |
| 1991 | 5.2 | 33.6 | 43.5 | 12.8 | 3.9 | 0.8 | 0.3 | 0.0 |
| Total | 4.1 | 38.6 | 35.1 | 15.3 | 5.8 | 1.0 | 0.1 | 0.0 |
| European |  |  |  |  |  |  |  |  |
| 1968 | 21.6 | 60.3 | 15.2 | 2.7 | 0.3 | 0.0 | 0.0 | 0.0 |
| 1969 | 0.0 | 83.8 | 16.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1970 | 0.0 | 90.4 | 9.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1971 | 9.3 | 66.5 | 19.9 | 3.1 | 1.2 | 0.0 | 0.0 | 0.0 |
| 1972 | 11.0 | 71.2 | 16.7 | 1.0 | 0.1 | 0.0 | 0.0 | 0.0 |
| 1973 | 26.0 | 58.0 | 14.0 | 2.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1974 | 22.9 | 68.2 | 8.5 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1975 | 26.0 | 53.4 | 18.2 | 2.5 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1976 | 23.5 | 67.2 | 8.4 | 0.6 | 0.3 | 0.0 | 0.0 | 0.0 |
| 1977 | - | - | - | 0. | - | - | - | 0.0 |
| 1978 | 26.2 | 64.5 | 8.2 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1979 | 23.6 | 64.8 | 11.0 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1980 | 25.8 | 56.9 | 14.7 | 2.5 | 0.2 | 0.0 | 0.0 | 0.0 |
| 1981 | 15.4 | 67.3 | 15.7 | 1.6 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1982 | 15.6 | 56.1 | 23.5 | 4.2 | 0.7 | 0.1 | 0.1 | 0.2 |
| 1983 | 34.7 | 50.2 | 12.3 | 2.4 | 0.3 | 0.2 | 0.0 | 0.0 |
| 1984 | 22.7 | 56.9 | 15.2 | 4.2 | 0.9 | 0.0 | 0.0 | 0.0 |
| 1985 | 20.2 | 61.6 | 14.9 | 2.7 | 0.6 | 0.0 | 0.1 | 0.0 |
| 1986 | 19.5 | 62.5 | 15.1 | 2.7 | 0.2 | 0.0 | 0.0 | 0.0 |
| 1987 | 19.2 | 62.5 | 14.8 | 3.3 | 0.3 | 0.0 | 0.0 | 0.0 |
| 1988 | 18.4 | 61.6 | 17.3 | 2.3 | 0.5 | 0.0 | 0.1 | 0.0 |
| 1989 | 18.0 | 61.7 | 17.4 | 2.7 | 0.3 | 0.0 | 0.0 | 0.0 |
| 1990 | 15.9 | 56.3 | 23.0 | 4.4 | 0.2 | 0.0 | 0.0 | 0.0 |
| 1991 | 20.9 | 47.4 | 26.3 | 4.2 | 1.2 | 0.0 | 0.0 | 0.0 |
| Total | 20.7 | 61.3 | 15.3 | 2.4 | 0.3 | 0.0 | 0.0 | 0.0 |

Table 5.1.3.1 Estimated Carlin tag recoveries and run size in Maine rivers. Ratio of tag to run size of 2 SW salmon in homewaters. Ratio (year i) for use in estimation of distant water harvest (year i-1).

| Year | Tags | Run | RATIO |
| :--- | ---: | ---: | ---: |
| 1967 | 0 | 1019 | 0.0000 |
| 1968 | 168 | 729 | 0.2307 |
| 1969 | 7 | 690 | 0.0104 |
| 1970 | 13 | 856 | 0.0155 |
| 1971 | 68 | 687 | 0.0985 |
| 1972 | 318 | 1449 | 0.2197 |
| 1973 | 206 | 1448 | 0.1425 |
| 1974 | 215 | 1411 | 0.1520 |
| 1975 | 450 | 2345 | 0.1920 |
| 1976 | 184 | 1341 | 0.1374 |
| 1977 | 97 | 2025 | 0.0478 |
| 1978 | 97 | 4145 | 0.0233 |
| 1979 | 36 | 1878 | 0.0190 |
| 1980 | 0 | 5662 | 0.0000 |
| 1981 | 470 | 5122 | 0.0918 |
| 1982 | 284 | 6023 | 0.0472 |
| 1983 | 138 | 1930 | 0.0716 |
| 1984 | 61 | 3045 | 0.0202 |
| 1985 | 185 | 4855 | 0.0381 |
| 1986 | 309 | 5568 | 0.0555 |
| 1987 | 119 | 2397 | 0.0498 |
| 1988 | 319 | 2855 | 0.1118 |
| 1989 | 190 | 2946 | 0.0646 |
| 1990 | 172 | 4370 | 0.0393 |
| 1991 | 29 | 2057 | 0.0138 |
|  |  |  |  |
|  |  |  |  |

Table 5.1.3.2. Carlin tag returns from 1SW salmon of Maine origin in West Greenland by year and NAFO division. $(99=$ NAFO division unknown)

| YEAR | 1 A | 1B | 1C | 1D | 1E | 1F | 99 | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 1 | 10 | 10 | 8 | 3 | 2 | 3 | 37 |
| 1969 | 0 | 1 | 3 | 0 | 1 | 0 | 1 | 6 |
| 1970 | 10 | 14 | 6 | 7 | 12 | 2 | 7 | 58 |
| 1971 | 29 | 34 | 50 | 57 | 58 | 60 | 94 | 382 |
| 1972 | 5 | 4 | 35 | 6 | 15 | 5 | 12 | 82 |
| 1973 | 5 | 28 | 25 | 16 | 13 | 12 | 32 | 131 |
| 1974 | 8 | 75 | 95 | 79 | 32 | 20 | 48 | 357 |
| 1975 | 10 | 22 | 16 | 5 | 1 | 3 | 70 | 127 |
| 1976 | 13 | 11 | 9 | 3 | 0 | 0 | 3 | 39 |
| 1977 | 0 | 1 | 6 | 0 | 1 | 2 | 1 | 11 |
| 1978 | 0 | 5 | 2 | 0 | 0 | 0 | 2 | 9 |
| 1980 | 0 | 37 | 20 | 9 | 0 | 0 | 6 | 72 |
| 1981 | 0 | 17 | 5 | 0 | 0 | 0 | 18 | 40 |
| 1982 | 1 | 42 | 1 | 1 | 0 | 2 | 2 | 49 |
| 1983 | 0 | 1 | 6 | 0 | 0 | 0 | 0 | 7 |
| 1984 | 1 | 9 | 9 | 0 | 1 | 3 | 0 | 23 |
| 1985 | 4 | 25 | 7 | 8 | 0 | 5 | 9 | 58 |
| 1986 | 1 | 10 | 15 | 17 | 11 | 18 | 0 | 72 |
| 1987 | 2 | 30 | 52 | 43 | 29 | 10 | 0 | 166 |
| 1988 | 1 | 29 | 24 | 28 | 20 | 4 | 0 | 106 |
| 1989 | 4 | 14 | 44 | 22 | 14 | 8 | 0 | 106 |
| 1990 | 1 | 2 | 6 | 4 | 2 | 0 | 0 | 15 |
| 1991 | 0 | 1 | 10 | 1 | 0 | 1 | 0 | 13 |
| Unk | 2 | 16 | 39 | 8 | 14 | 3 | 5 | 87 |
| TOTAL | 98 | 438 | 495 | 322 | 227 | 160 | 313 | 2053 |

Table 5.1.3.3. Estimated harvest of 1SW salmon of Maine origin in West Greenland by year and NAFO division. ( $99=$ NAFO division unknown)

| YEAR | 1 A | 1B | 1 C | 1D | $1 E$ | 1F | 99 | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1967 | 6 | 61 | 61 | 49 | 18 | 12 | 18 | 226 |
| 1969 | 0 | 91 | 273 | 0 | 91 | 0 | 91 | 545 |
| 1970 | 143 | 200 | 86 | 100 | 171 | 29 | 100 | 828 |
| 1971 | 186 | 218 | 320 | 365 | 371 | 384 | 602 | 2446 |
| 1972 | 49 | 39 | 345 | 59 | 148 | 49 | 118 | 809 |
| 1973 | 46 | 259 | 231 | 148 | 120 | 111 | 296 | 1212 |
| 1974 | 59 | 549 | 696 | 579 | 234 | 147 | 352 | 2615 |
| 1975 | 102 | 225 | 164 | 51 | 10 | 31 | 716 | 1299 |
| 1976 | 510 | 431 | 353 | 118 | 0 | 0 | 118 | 1529 |
| 1977 | 0 | 81 | 483 | 0 | 81 | 161 | 81 | 886 |
| 1978 | 0 | 592 | 237 | 0 | 0 | 0 | 237 | 1066 |
| 1980 | 0 | 1134 | 613 | 276 | 0 | 0 | 184 | 2207 |
| 1981 | 0 | 811 | 238 | 0 | 0 | 0 | 858 | 1908 |
| 1982 | 26 | 1100 | 26 | 26 | 0 | 52 | 52 | 1283 |
| 1983 | 0 | 70 | 418 | 0 | 0 | 0 | 0 | 488 |
| 1984 | 37 | 332 | 332 | 0 | 37 | 111 | 0 | 849 |
| 1985 | 101 | 633 | 177 | 203 | 0 | 127 | 228 | 1469 |
| 1986 | 28 | 283 | 424 | 480 | 311 | 509 | 0 | 2035 |
| 1987 | 25 | 377 | 654 | 541 | 365 | 126 | 0 | 2087 |
| 1988 | 22 | 632 | 523 | 610 | 436 | 87 | 0 | 2309 |
| 1989 | 143 | 501 | 1576 | 788 | 501 | 287 | 0 | 3797 |
| 1990 | 102 | 203 | 610 | 407 | 203 | 0 | 0 | 1525 |
| TOTAL | 1585 | 8823 | 8841 | 4799 | 3098 | 2221 | 4051 | 33418 |

Table 5.1.3.4. The estimated number of CWTs and total number of fish in the 2SW run, and the RATIO used to raise tag recoveries to harvest estimates.

|  | Maine |  |  | Merrimack |  |  | Connecticut |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | CWT | Run | RATIO | CWT | Run | RATIO | CWT | Run | RATIO |
| 1988 | 603 | 2855 | 0.211 | 13 | 54 | 0.241 | 93 | 94 | 0.989 |
| 1989 | 634 | 2946 | 0.215 | 3 | 79 | 0.038 | 58 | 106 | 0.547 |
| 1990 | 456 | 4370 | 0.104 | 110 | 219 | 0.502 | 226 | 262 | 0.863 |
| 1991 | 438 | 2057 | 0.213 | 69 | 329 | 0.210 | 153 | 202 | 0.757 |

Table 5.1.3.5. CWT harvest estimates for USA stocks. Harvest is in numbers and CV is coefficient of variation.

|  | MAINE |  | MERRIMACK |  | CONNECTICUT |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| YEAR | HARVEST | CV | HARVEST | CV | HARVEST | CV |
| 1987 | 5571 | 10 | 51 | 51 | 51 | 69 |
| 1988 | 3882 | 13 |  |  | 76 | 55 |
| 1989 | 2857 | 13 | 243 | 21 | 115 | 22 |
| 1990 | 1613 | 26 | 1070 | 35 | 231 | 39 |

Table 5.1.4.1. Distributions of recovery rates (per 1000 fish examined) of microtagged salmon from different countries caught in the West Greenland fishery, 1986-1991.


Notes: Empty cells represent no scanning.
Years in which no tags were recovered from a particular
country were not included.

Table 5.1.4.2 Trend analysis for North American proportions at Greenland

| Year | TEMPORAL |  |  |  |  | SPATIAL |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | NAFO | Div | R squared |  | Sig. | NAFO | Div | Z-Value |  | Sig. |
| 1990 |  | 1B | 0.87 |  | NS | 1 C vs | 1D | -0.02 |  | NS |
|  |  | 1D | 0.29 |  | NS | 1 C vs | 1E | -5.19 | $<$ | 0.01 |
|  |  |  |  |  |  | 1 D vs | 1E | -5.16 | < | 0.01 |
| 1989 |  | 1B | 0.02 |  | NS | 1 B vs | 1D | -2.35 | $<$ | 0.01 |
|  |  | 1D | 0.01 |  | NS | 1B vs | 1E | -7.83 | < | 0.01 |
|  |  | 1E | 0.01 |  | NS | 1D vs | 1E | -2.36 | < | 0.01 |
| 1988 |  | 1B | 0.07 |  | NS | 1 A vs | 1 B | -1.41 |  | NS |
|  |  | 1D | 0.4 | $<$ | 0.05 | 1 A vs | 1D | -3.10 | < | 0.05 |
|  |  | 1 E | 0.19 |  | NS | 1 A vs | 1E | -1.50 |  | NS |
|  |  |  |  |  |  | 1 A vs | 1 F | -0.50 |  | NS |
|  |  |  |  |  |  | 1 B vs | 1D | -2.10 | < | 0.05 |
|  |  |  |  |  |  | 1 B vs | 1 E | -0.10 |  | NS |
|  |  |  |  |  |  | 1 B vs | $1 F$ | 0.80 |  | NS |
|  |  |  |  |  |  | 1D vs | 1 E | 2.10 | $<$ | 0.05 |
|  |  |  |  |  |  | 1D vs | 1F | 2.40 | < | 0.05 |
|  |  |  |  |  |  | 1 E vs | 1F | 0.80 |  | NS |
| 1987 |  | 1B | 0.72 | < | 0.05 | 1 B vs | 1D | - |  | NS |
|  |  | 1D | 0.02 |  | NS | 1 B vs | 1 E | - | $<$ | 0.05 |
|  |  | 1 E | 0.31 |  | NS | 1 B vs | 1 F | - | $<$ | 0.05 |
|  |  | 1 F | 0.15 |  | NS | 1 D vs | 1 E | - | $<$ | 0.05 |
|  |  |  |  |  |  | 1 D vs | 1 F | - | $<$ | 0.05 |
|  |  |  |  |  |  | 1 E vs | 1 F | - | < | 0.05 |
| 1986 |  | 1B | 0.44 |  | NS | 1 B vs | 1D | - |  | NS |
|  |  | 1D | 0.18 |  | NS | 1 B vs | 1 E | - |  | NS |
|  |  | 1 E | 0.06 |  | NS | 1D vs | 1E | - |  | NS |

Table 5.1.4.3 Results of ANOVA on catches on North American Proportion of catch, 1983 - 1990.

| Dependant Variable: ARCNA |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Source | DF | Sum of <br> Squares | Mean <br> Square | F Value | Prob. |
| Model | 33 | 1.53744 | 0.04659 | 3.43 | 0.0001 |
| Error | 154 | 2.08882 | 0.01357 |  |  |
| Corrected <br> Total | 187 | 3.62627 |  |  |  |


| R-Squared | C.V. | Root MSE | ARCNA Mean |
| :--- | :--- | :---: | :---: |
| 0.423974 | 20.89814 | 0.116464 | 0.55729 |

Dependant Variable: ARCNA

| Source | DF | Type III SS | Mean Square | F Value | Prob. |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  |  |  |  |  |
| YEAR | 7 | 0.88756 | 0.12679 | 9.35 | 0.0001 |
| NAFO | 5 | 0.12785 | 0.02557 | 1.89 | 0.1000 |
| WEEK | 7 | 0.04000 | 0.00571 | 0.42 | 0.8879 |
| YEAR*NAFO | 14 | 0.42298 | 0.03021 | 2.23 | 0.0090 |

Table 5.3.1.1. Definition of key variables used in continental run-reconstruction model for North America.

| Variable | Definition |
| :---: | :---: |
| i | Year of the fishery on 1SW salmon in Greenland and Canada. |
| i+1 | Year of return to Canada for 2SW salmon. |
| T2(i+1) | Total population size of 2SW salmon prior to fishery in home waters in year i+1 |
| Hc2( $\mathrm{i}+1$ ) | Harvest of 2SW salmon in Canada in year i+1 |
| $\mathrm{Hc} 2(\mathrm{i})$ | Harvest of 2SW salmon in Canada in year i |
| Hg1 (i) | Harvest of 1SW salmon in Greenland in year i |
| P | fraction of extant population available to fishery in Canada |
| FO | fraction of population not available to any fishery (e.g., Labrador Sea or Irminger Sea |
| 1-P-FU | fraction of extant population available to fishery in Greenland in year i |
| Ec(i) | exploitation rate on 1SW salmon in Canada in year i |
| R2(i+1) | total 2SW run to rivers in Canada in year i+1 |
| R1(i) | total $15 W$ run to rivers in Canada in year i |
| \% | Natural mortality rate (.12/yr). |
| t1 | time between the midpoint of the Canadian fishery and return to river. $=2$ months |
| t2 | time between the midpoint of Greenland fishery and return to Canada $=10$ months. |
| S1 | Survival of 2 SW salmon between the homewater fishery and return to river $\{\exp (-\mathrm{Mt} \mathrm{t})$ \} |
| S2 | Survival of 1SW salmon between the Greenland fishery and 2SW homewater fishery. $\{\exp (-\mathrm{M}$ t2) $\}$ |
| H_S(i) | Number of "Small" salmon caught in Canada in year i; fish $<2.7 \mathrm{~kg}$ |
| H_L(i) | Number of "Large" salmon caught in Canada in year i; fish $>2.7 \mathrm{~kg}$ |
| f_imm | Fraction of 1SW salmon that are immature, i.e., non-maturing; range $=0.1$ to 0.2 |
| q | Fraction of 1SW salmon present in the Large size market category; range $=0.1$ to 0.3 |

Table 5.3.1.2
a) Summary of catch and returns (numbers) by size class (Small, Large) for Canada. Partitioning of catches by Salmon Fishing Area (SFA) is related to known migration patterns of 1SW salmon and to availability of grilse return estimates.

| Year$\qquad$ | $\begin{array}{\|c\|} \{\text { SFA }\} \\ \hline \begin{array}{c} \text { Grld Catch } \\ \text { (i) } \\ \hline \end{array}{ }^{2} \\ \hline \end{array}$ | NonMat 1SW Component$\{1-7,14 b\}$ |  | Mat 1SW Component$\{3-7,14 a\}$ |  | Grilse Returns $\{3-7,14 a\}$ |  | 2SW Catches |  | Total 2SW Returns <br> All Canada |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | \{1-7, 14b $\}$ | \{8-14a\} |  |  |  |  |
|  |  | H_Small <br> (i) | $\begin{gathered} \text { H_Large } \\ \text { (i) } \\ \hline \end{gathered}$ |  |  | H_Small $\qquad$ | H_Large (i) | Min <br> (i) | $\begin{gathered} \hline \text { Max } \\ \text { (i) } \\ \hline \end{gathered}$ | $\begin{array}{r} \text { H_Large } \\ (i+1) \end{array}$ | $\begin{array}{r} \text { H_Large } \\ (i+1) \\ \hline \end{array}$ | $\begin{gathered} \operatorname{Min} \\ (i+1) \end{gathered}$ | $\begin{gathered} \text { Max } \\ (i+1) \end{gathered}$ |
| 1983 | 33240 | 166033 | 100965 | 132705 | 57663 | 57210 | 85815 | 87742 | 15135 | 88434 | 146349 |
| 1984 | 38910 | 123674 | 87742 | 104118 | 44540 | 62730 | 94095 | 70970 | 24383 | 97074 | 166799 |
| 1985 | 139230 | 178719 | 70970 | 140635 | 36975 | 70677 | 106015 | 107561 | 22036 | 122089 | 209043 |
| 1986 | 171740 | 222641 | 107561 | 164490 | 48996 | 65707 | 98560 | 146242 | 19241 | 105162 | 177538 |
| 1987 | 173680 | 281771 | 146242 | 217558 | 67072 | 41550 | 62325 | 86047 | 14763 | 117132 | 199477 |
| 1988 | 116760 | 198484 | 86047 | 138316 | 36449 | 75280 | 112920 | 85319 | 15577 | 103710 | 178279 |
| 1989 | 60690 | 172861 | 85319 | 128838 | 37576 | 31520 | 47280 | 59334 | 11639 | 103664 | 186349 |
| 1990 | 61710 | 104788 | 59334 | 80383 | 31847 | 50753 | 76130 | 38450 | 9155 | 93453 | 156633 |
|  | [G1] |  |  |  |  | [R1_min] | __max] | [H_L(1-7)] | [H_L(8-14)] | [R2_min] | _max] |

b) Summary of estimated catch and return data by sea age class for Canada. See text and designated equations for details on computation.

| $\begin{array}{\|r} \text { Year } \\ \text { (i) } \\ \hline \end{array}$ | \{SFA 1-7, 14a\} |  | \{SFA 3-7, 14a\} |  | \{SFA 3-7, 14a\} |  | \{SFA 1-13, 14a, 14b\} |  | \{SFA 1-13, 14a, 14b \} |  | $\begin{array}{\|c\|} \hline \text { 2SW Catch } \\ \hline \text { in River } \\ \hline \end{array}$ | Pre-fishery nonmaturing 1SW abundance (N. Amer) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW Non Maturing (i) |  | 1SW Maturing (i) |  | Grilse Returns (i) |  | 2SW Harvest (i+1) |  | 2SW Returns ( $i+1$ ) |  |  |  |  |
|  | min | max | min | max | min | max | min | $\max$ | min | max | (i+1) | min | max |
| 1983 | 17613 | 39265 | 110777 | 135004 | 57210 | 85815 | 76554 | 94103 | 88434 | 146349 | 25149 | 234176 | 339870 |
| 1984 | 13245 | 29999 | 86858 | 105732 | 62730 | 94095 | 74062 | 88256 | 97074 | 166799 | 27302 | 242368 | 352641 |
| 1985 | 18582 | 40002 | 115466 | 136555 | 70677 | 106015 | 97329 | 118841 | 122089 | 209043 | 34092 | 401662 | 543922 |
| 1986 | 23340 | 50982 | 135512 | 161270 | 65707 | 98560 | 121610 | 150859 | 105162 | 177538 | 33061 | 446870 | 587628 |
| 1987 | 29640 | 65129 | 179412 | 213912 | 41550 | 62325 | 74996 | 92205 | 117132 | 199477 | 32417 | 416955 | 563383 |
| 1988 | 20709 | 44860 | 113569 | 134326 | 75280 | 112920 | 75300 | 92364 | 103710 | 178279 | 27583 | 336458 | 462707 |
| 1989 | 18139 | 39691 | 106076 | 126100 | 31520 | 47280 | 53173 | 65040 | 103664 | 186349 | 26026 | 253312 | 380279 |
| 1990 | 11072 | 24518 | 66854 | 80943 | 50753 | 76130 | 36070 | 43760 | 93453 | 156633 | 26056 | 216965 | 309436 |
|  | \{Eq.7\} | \{Eq.6\} | \{Eq.23\} | \{Eq.22\} |  |  | \{Eq.16\} | Eq.17\} |  |  |  |  |  |

$\left[\mathrm{N} 1 \_\right.$min $]=\left(\left[\mathrm{R} 2 \_\right.\right.$min $\left.] / \mathrm{S} 1+\left[\mathrm{C} 2 \_m i n\right]\right) / \mathrm{S} 2+\left[\mathrm{C} 1 \_\right.$min $]+[\mathrm{G} 1]$
$\left[\mathrm{N} 1 \_\max \right]=\left(\left[\mathrm{R} 2_{-} \max \right] / \mathrm{S} 1+\left[\mathrm{C} 2 \_\max \right]\right) / \mathrm{S} 2+\left[\mathrm{C}_{1} \max \right]+[\mathrm{G} 1]$

$$
\text { where } S 1=\exp \left(-0.12^{\star} 2 / 12\right) \text { and } S 2=\exp \left(-0.12^{\star} 10 / 12\right)
$$

Table 5.3.1.3. Exploitation rates and population abundance estimates of salmon at West Greenland; derived from constraints model.

| Year(i) | Fract. <br> Unavail (FU) | Exploitation |  |  |  | Proportion of Stock in Canada |  | W. Grld <br> Catch <br> HG1 | Prop <br> N. Amer <br> fna | Estimated Range of Abundance of Salmon at West Greenland |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Canada |  | Greenland |  |  |  | Total Population |  | N. American Stocks |  | European Stocks |  |
|  |  | Ec_min | Ec_max | Eg_min | Eg_max | P_min | P_max |  |  | Ng _min | $\mathrm{Ng}_{\text {_ }}$ max | Nna_min | Nna_max | Neu_min | Neu_max |
| 1983 | 0.05 | 56 | 70 | 12 | 19 | 6 | 14 |  | 100000 | 40 | 526316 | 833333 | 210526 | 333333 | 315789 | 500000 |
| 1983 | 0.3 | 56 | 70 | 16 | 25 | 6 | 13 | 100000 | 40 | 400000 | 625000 | 160000 | 250000 | 240000 | 375000 |
| 1984 | 0.05 | 48 | 63 | 13 | 21 | 5 | 13 | 95500 | 50 | 454762 | 734615 | 227381 | 367308 | 227381 | 367308 |
| 1984 | 0.3 | 48 | 63 | 17 | 28 | 5 | 12 | 95500 | 50 | 341071 | 561765 | 170536 | 280882 | 170536 | 280882 |
| 1985 | 0.05 | 52 | 66 | 29 | 42 | 5 | 12 | 301050 | 50 | 716786 | 1038103 | 358393 | 519052 | 358393 | 519052 |
| 1985 | 0.3 | 52 | 66 | 36 | 51 | 5 | 11 | 301050 | 50 | 590294 | 836250 | 295147 | 418125 | 295147 | 418125 |
| 1986 | 0.05 | 58 | 71 | 34 | 46 | 2 | 11 | 316830 | 57 | 688761 | 931853 | 392594 | 531156 | 296167 | 400697 |
| 1986 | 0.3 | 58 | 71 | 41 | 55 | 3 | 11 | 316830 | 57 | 576055 | 772756 | 328351 | 440471 | 247703 | 332285 |
| 1987 | 0.05 | 74 | 84 | 37 | 49 | 2 | 9 | 305700 | 59 | 623878 | 826216 | 368088 | 487468 | 255790 | 338749 |
| 1987 | 0.3 | 74 | 84 | 44 | 58 | 2 | 8 | 305700 | 59 | 527069 | 694773 | 310971 | 409916 | 216098 | 284857 |
| 1988 | 0.05 | 50 | 64 | 30 | 45 | 4 | 17 | 280820 | 43 | 624044 | 936067 | 268339 | 402509 | 355705 | 533558 |
| 1988 | 0.3 | 50 | 64 | 37 | 55 | 3 | 17 | 280820 | 43 | 510582 | 758973 | 219550 | 326358 | 291032 | 432615 |
| 1989 | 0.05 | 69 | 80 | 19 | 29 | 2 | 8 | 117420 | 56 | 404897 | 618000 | 226742 | 346080 | 178154 | 271920 |
| 1989 | 0.3 | 69 | 80 | 24 | 37 | 2 | 7 | 117420 | 56 | 317351 | 489250 | 177717 | 273980 | 139635 | 215270 |
| 1990 | 0.05 | 47 | 62 | 23 | 36 | 3 | 14 | 85980 | 75 | 238833 | 373826 | 179125 | 280370 | 59708 | 93457 |
| 1990 | 0.3 | 47 | 62 | 29 | 46 | 3 | 14 | 85980 | 75 | 186913 | 296483 | 140185 | 222362 | 46728 | 74121 |

Table 5.3.1.4. Estimation of the number of potential 2SW salmon in the sea after the West Greenland fishery in 1989

| Country | 1990 <br> $H m$ | q | E <br> $(\mathrm{min})$ | E <br> $(\max )$ | t | S 1 | T 1 g <br> $(\mathrm{~min})$ <br> $(000 \mathrm{~s})$ | T 1 g <br> $(\mathrm{max})$ <br> $(000 \mathrm{~s})$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| S.Europe: |  |  |  |  |  |  |  |  |
|         <br> UK(Sc) 54416 0.8 0.25 0.40 8 0.92 179 |  |  |  |  |  |  |  |  |
| UK(E\&W) | 33533 | 1 | 0.30 | 0.50 | 8 | 0.92 | 110 | 183 |
| UK(NI) | 30745 | 0.1 | 0.40 | 0.80 | 8 | 0.92 | 6 | 13 |
| Ireland | 30000 | 0.95 | 0.30 | 0.40 | 8 | 0.92 | 117 | 156 |
| France | 2186 | 1 | 0.40 | 0.50 | 8 | 0.92 | 7 | 9 |

N.Europe:

| Norway | 92759 | 0.7 | 0.60 | 0.70 | 10 | 0.90 | 155 | 181 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Sweden | 3133 | 0.8 | 0.70 | 0.90 | 10 | 0.90 | 5 | 6 |
| Finland | 5240 | 0.7 | 0.60 | 0.80 | 10 | 0.90 | 8 | 10 |
| Russia | 20633 | 1 | 0.55 | 0.65 | 10 | 0.90 | 53 | 63 |


| Iceland | 18253 | 1 | 0.25 | 0.50 | 8 | 0.92 | 60 | 120 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | | Faroes | 103407 | 1 | 1.00 | 1.00 | 4 | 0.96 | 163 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |


| Extant stock of N.Europe (ex F. catch + Iceland) $=$ | 221 | 260 |  |  |
| :--- | ---: | ---: | :---: | :---: |
| Extant stock of S.Europe (ex F. catch + Iceland) $=$ | 419 | 647 |  |  |
|  |  |  |  |  |


| Harvest of 2SW European fish at WG $=$ | 52 |
| :--- | :--- |


| Total number of 1SW European salmon in the sea <br> before the West Greenland fishery $=$ | 915 |  |
| :--- | ---: | ---: |


| $M=$ | 0.01 |
| :---: | :---: |
| $R=$ | 0.34 |

Table 5.3.3.1. Variability in estimates of the proportion of salmon maturing at 1SW and post-smolt survival for monitored rivers in the North east Atlantic using the run reconstruction model.

Pm1 - Proportion maturing at 1SW

| Year | R.Imsa |  |  | R.Drammen |  | N.Esk | R.Bush |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | W | H1+ | H2+ | H1+ | H2+ | W | H |
| 1980 |  |  |  |  |  |  |  |
| 1981 | 0.68 | 0.50 | 0.76 |  |  | 0.64 |  |
| 1982 | 0.59 | 0.83 | 0.73 |  |  | 0.74 |  |
| 1983 | 0.81 | 0.48 | 0.86 |  |  | 0.73 |  |
| 1984 | 0.65 | 0.68 | 0.84 | 0.23 | 0.45 | 0.64 |  |
| 1985 | 0.63 | 0.78 | 0.84 | 0.39 |  | 0.71 | 0.97 |
| 1986 | 0.36 | 0.84 | 0.83 | 0.62 |  | 0.52 | 0.97 |
| 1987 | 0.65 | 0.85 | 0.87 | 0.33 | 0.55 | 0.71 | 0.97 |
| 1988 | 0.88 | 0.63 | 0.96 | 0.38 |  | 0.79 | 0.97 |
| 1989 | 0.59 |  |  |  | 0.57 | 0.69 | 0.97 |
| 1990 |  |  |  |  |  |  |  |
| Mean | 0.65 | 0.70 | 0.84 | 0.39 | 0.44 | 0.69 | 0.97 |
| S.D. | 0.16 | 0.15 | 0.07 | 0.14 | 0.13 | 0.07 | 0.00 |

PS - Post smolt survival

| Year | R.Imsa |  |  | R.Drammen |  | $\begin{array}{\|c} \text { N.Esk } \\ \hline W \end{array}$ | $\begin{array}{\|r} \hline \text { R.Bush } \\ \hline H \end{array}$ | $\begin{array}{\|c\|} \hline \text { Burrishoole } \\ \hline \mathrm{H} \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | W | H1+ | H2+ | H1+ | H2+ |  |  |  |
| 1980 |  |  |  |  |  |  |  | 0.04 |
| 1981 | 0.31 | 0.03 | 0.16 |  |  | 0.06 |  | 0.09 |
| 1982 | 0.11 | 0.01 | 0.11 |  |  | 0.09 |  | 0.08 |
| 1983 | 0.20 | 0.03 | 0.04 |  |  | 0.04 | 0.06 | 0.03 |
| 1984 | 0.16 | 0.11 | 0.07 | 0.17 | 0.13 | 0.04 | 0.15 | 0.17 |
| 1985 | 0.20 | 0.08 | 0.06 | 0.11 |  | 0.05 | 0.18 | 0.26 |
| 1986 | 0.13 | 0.16 | 0.07 | 0.12 |  | 0.06 | 0.10 | 0.07 |
| 1987 | 0.33 | 0.14 | 0.12 | 0.02 | 0.05 | 0.03 | 0.12 | 0.15 |
| 1988 | 0.19 | 0.06 | 0.13 | 0.02 |  | 0.02 | 0.12 | 0.19 |
| 1989 | 0.12 |  |  |  | 0.04 | 0.03 | 0.11 | 0.06 |
| 1990 |  |  |  |  |  |  | 0.06 | 0.11 |
| Mean | 0.19 | 0.08 | 0.10 | 0.09 | 0.08 | 0.05 | 0.11 | 0.11 |
| S.D. | 0.07 | 0.05 | 0.04 | 0.07 | 0.06 | 0.02 | 0.04 | 0.07 |

$H=$ Hatchery reared
$W=$ Wild

Table 5.3.4.1. Summary of input data for estimation of exploitation rates for Maine origin Atlantic salmon.

| Year <br> $(i)$ | Run2 <br> $(i+1)$ | GH2 <br> $(i+1)$ | CH2 <br> $(i+1)$ | USAC <br> $(i+1)$ | NN2 <br> $(i+1)$ | GH1 <br> $(i)$ | CH1 <br> $(i)$ | RUN3 <br> $(i+2)$ | RUN1 <br> $(i)$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1967 | 729 | 18 | 50 | 0 | 161 | 226 | 242 | 19 | 24 |
| 1968 | 690 | 135 | 274 | 0 | 274 | 0 | 411 | 18 | 36 |
| 1969 | 856 | 0 | 92 | 0 | 92 | 545 | 277 | 17 | 14 |
| 1970 | 687 | 100 | 135 | 14 | 14 | 828 | 398 | 49 | 44 |
| 1971 | 1449 | 77 | 12 | 7 | 52 | 2446 | 295 | 13 | 32 |
| 1972 | 1448 | 118 | 66 | 30 | 20 | 809 | 105 | 59 | 43 |
| 1973 | 1411 | 65 | 9 | 28 | 38 | 1212 | 220 | 28 | 99 |
| 1974 | 2345 | 73 | 65 | 30 | 7 | 2615 | 755 | 5 | 116 |
| 1975 | 1341 | 0 | 8 | 0 | 0 | 1299 | 1014 | 16 | 231 |
| 1976 | 2025 | 0 | 90 | 30 | 60 | 1529 | 2230 | 32 | 98 |
| 1977 | 4110 | 80 | 61 | 0 | 0 | 878 | 933 | 4 | 161 |
| 1978 | 1878 | 0 | 59 | 0 | 0 | 1066 | 309 | 33 | 847 |
| 1980 | 5107 | 61 | 135 | 0 | 0 | 2200 | 4607 | 41 | 1148 |
| 1981 | 6003 | 143 | 144 | 30 | 60 | 1901 | 1137 | 15 | 315 |
| 1982 | 1915 | 104 | 31 | 0 | 20 | 1273 | 1586 | 16 | 271 |
| 1983 | 3025 | 69 | 0 | 0 | 0 | 485 | 1689 | 8 | 388 |
| 1984 | 4830 | 0 | 95 | 0 | 0 | 844 | 1322 | 24 | 337 |
| 1985 | 5563 | 51 | 66 | 0 | 0 | 1468 | 2274 | 52 | 711 |
| 1986 | 2397 | 0 | 0 | 0 | 0 | 2035 | 533 | 7 | 950 |
| 1987 | 2870 | 38 | 49 | 13 | 0 | 2086 | 584 | 8 | 896 |
| 1988 | 2946 | 22 | 61 | 44 | 0 | 2308 | 393 | 21 | 1267 |
| 1989 | 4370 | 36 | 28 | 0 | 0 | 3797 | 1722 | 5 | 654 |
| 1990 | 2057 | 0 | 103 | 0 | 0 | 1525 | 780 | 0 | 301 |

KEY Run2 $=$ Estimated total run of 2SW salmon to Maine rivers
$\mathrm{GH} 2=$ Harvest of 2SW salmon in Greenland
CH2 = Harvest of 2SW salmon in Canada
USAC = Harvest of 2SW salmon in USA coastal waters
NN2 = Non- Newfoundland 2SW harvests
GH1 = Harvest of 1SW salmon in Greenland
$\mathrm{CH} 1=$ Harvest of 1SW salmon in Canada
RUN3 = Estimated total run of 3SW salmon to Maine Rivers
RUN1 = Estimated total run of 1SW salmon to Maine Rivers

Table 5.3.4.2. Estimated exploitation rate of 1 SW salmon for the extant population of Maine origin stocks. Only columns with 0.00 unaccounted for are true "extant" exploitation rates.

| Natural Mortality | 0.12 | 0.12 | 0.12 | 0.12 | 0.24 | 0.24 | 0.24 | 0.24 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Fraction Unaccounted | 0 | 0 | 0.1 | 0.1 | 0 | 0 | 0.1 | 0.1 |
| Adjusted Carlin | 1 | 2 | 1 | 2 | 1 | 2 | 1 | 2 |


| 1967 | 0.30 | 0.41 | 0.32 | 0.43 | 0.28 | 0.38 | 0.30 | 0.41 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1968 | 0.21 | 0.26 | 0.23 | 0.28 | 0.19 | 0.24 | 0.21 | 0.26 |
| 1969 | 0.41 | 0.54 | 0.44 | 0.57 | 0.38 | 0.51 | 0.41 | 0.54 |
| 1970 | 0.52 | 0.63 | 0.55 | 0.66 | 0.49 | 0.60 | 0.52 | 0.63 |
| 1971 | 0.60 | 0.74 | 0.63 | 0.76 | 0.58 | 0.71 | 0.60 | 0.73 |
| 1972 | 0.32 | 0.45 | 0.34 | 0.48 | 0.29 | 0.42 | 0.32 | 0.45 |
| 1973 | 0.45 | 0.60 | 0.47 | 0.62 | 0.42 | 0.57 | 0.45 | 0.60 |
| 1974 | 0.54 | 0.69 | 0.57 | 0.71 | 0.52 | 0.67 | 0.54 | 0.69 |
| 1975 | 0.60 | 0.75 | 0.63 | 0.77 | 0.58 | 0.73 | 0.60 | 0.75 |
| 1976 | 0.60 | 0.74 | 0.63 | 0.76 | 0.57 | 0.71 | 0.60 | 0.73 |
| 1977 | 0.28 | 0.42 | 0.30 | 0.45 | 0.25 | 0.40 | 0.28 | 0.42 |
| 1978 | 0.38 | 0.55 | 0.41 | 0.57 | 0.36 | 0.52 | 0.38 | 0.55 |
| 1980 | 0.53 | 0.69 | 0.56 | 0.71 | 0.51 | 0.66 | 0.53 | 0.69 |
| 1981 | 0.30 | 0.45 | 0.32 | 0.47 | 0.28 | 0.42 | 0.30 | 0.44 |
| 1982 | 0.55 | 0.70 | 0.58 | 0.72 | 0.52 | 0.67 | 0.55 | 0.69 |
| 1983 | 0.39 | 0.55 | 0.41 | 0.58 | 0.36 | 0.52 | 0.38 | 0.55 |
| 1984 | 0.28 | 0.43 | 0.30 | 0.46 | 0.26 | 0.41 | 0.28 | 0.43 |
| 1985 | 0.37 | 0.53 | 0.39 | 0.56 | 0.34 | 0.51 | 0.37 | 0.53 |
| 1986 | 0.49 | 0.66 | 0.52 | 0.68 | 0.46 | 0.63 | 0.49 | 0.66 |
| 1987 | 0.45 | 0.61 | 0.47 | 0.63 | 0.42 | 0.58 | 0.44 | 0.61 |
| 1988 | 0.44 | 0.60 | 0.47 | 0.63 | 0.41 | 0.57 | 0.44 | 0.60 |
| 1989 | 0.53 | 0.69 | 0.55 | 0.71 | 0.50 | 0.66 | 0.53 | 0.69 |
| 1990 | 0.49 | 0.65 | 0.52 | 0.67 | 0.46 | 0.62 | 0.49 | 0.65 |


| Average: Time Series | 0.43 | 0.59 | 0.45 | 0.61 | 0.40 | 0.56 | 0.43 | 0.59 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Average: Last Ten Years | 0.44 | 0.58 | 0.46 | 0.60 | 0.41 | 0.55 | 0.43 | 0.58 |

Table 5.3.4.3. Estimate exploitation rate of 2 SW salmon for the extant population of Maine origin stocks. Only columns with 0.00 unaccounted for are true "extant" exploitation rates.

| Natural Mortality | 0.12 | 0.12 | 0.12 | 0.12 | 0.24 | 0.24 | 0.24 | 0.24 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Fraction Unaccounted | 0 | 0 | 0.1 | 0.1 | 0 | 0 | 0.1 | 0.1 |
| Adjusted Carlin | 1 | 2 | 1 | 2 | 1 | 2 | 1 | 2 |


| 1968 | 0.76 | 0.87 | 0.78 | 0.88 | 0.74 | 0.85 | 0.76 | 0.86 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1969 | 0.95 | 0.98 | 0.96 | 0.98 | 0.95 | 0.97 | 0.95 | 0.98 |
| 1970 | 0.83 | 0.91 | 0.84 | 0.92 | 0.81 | 0.90 | 0.83 | 0.91 |
| 1971 | 0.81 | 0.90 | 0.83 | 0.91 | 0.79 | 0.89 | 0.81 | 0.90 |
| 1972 | 0.86 | 0.92 | 0.87 | 0.93 | 0.85 | 0.92 | 0.86 | 0.92 |
| 1973 | 0.74 | 0.85 | 0.76 | 0.86 | 0.72 | 0.83 | 0.74 | 0.85 |
| 1974 | 0.70 | 0.83 | 0.72 | 0.84 | 0.68 | 0.81 | 0.70 | 0.83 |
| 1975 | 0.96 | 0.98 | 0.96 | 0.98 | 0.96 | 0.98 | 0.96 | 0.98 |
| 1976 | 0.31 | 0.47 | 0.33 | 0.50 | 0.29 | 0.45 | 0.31 | 0.47 |
| 1977 | 0.72 | 0.83 | 0.74 | 0.85 | 0.69 | 0.82 | 0.72 | 0.83 |
| 1978 | 0.97 | 0.98 | 0.97 | 0.99 | 0.97 | 0.98 | 0.97 | 0.98 |
| 1980 | 0.62 | 0.76 | 0.64 | 0.78 | 0.59 | 0.74 | 0.62 | 0.76 |
| 1981 | 0.81 | 0.90 | 0.83 | 0.91 | 0.79 | 0.89 | 0.81 | 0.90 |
| 1982 | 0.94 | 0.97 | 0.95 | 0.97 | 0.94 | 0.97 | 0.94 | 0.97 |
| 1983 | 0.88 | 0.94 | 0.89 | 0.94 | 0.87 | 0.93 | 0.88 | 0.94 |
| 1984 | 0.89 | 0.94 | 0.90 | 0.95 | 0.87 | 0.93 | 0.89 | 0.94 |
| 1985 | 0.78 | 0.88 | 0.80 | 0.89 | 0.76 | 0.86 | 0.78 | 0.88 |
| 1986 | 0.67 | 0.80 | 0.69 | 0.82 | 0.64 | 0.78 | 0.67 | 0.80 |
| 1987 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1988 | 0.91 | 0.95 | 0.92 | 0.96 | 0.90 | 0.95 | 0.91 | 0.95 |
| 1989 | 0.78 | 0.88 | 0.80 | 0.89 | 0.76 | 0.86 | 0.78 | 0.88 |
| 1990 | 0.92 | 0.96 | 0.93 | 0.96 | 0.91 | 0.95 | 0.92 | 0.96 |


| Average: Time Series | 0.76 | 0.82 | 0.77 | 0.83 | 0.75 | 0.81 | 0.76 | 0.82 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Average: Last Ten Years | 0.76 | 0.84 | 0.78 | 0.85 | 0.75 | 0.83 | 0.76 | 0.84 |

Table 5.3.4.4. Estimates of exploitation rates for the reduced model in the fisheries of Newfoundland-Labrador and West Greenland for varying levels of $P$, the fraction of the stock migrating from Canada, (1-P, is fraction from Greenland) and for two levels of adjustment for reporting rate of Carlin tags.
The fraction of the stock unaccounted for is assumed to be 0.0 for these estimates.

|  | Carlin Adjustment $=1.0$ <br> Evaluation of P-fraction |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Year <br> Year | Can. | Grld. | Can. | Grld. | Can. | Grld. |
|  | 0.1 | 0.9 | 0.5 | 0.5 | 0.9 | 0.1 |
| 1967 | 0.75 | 0.24 | 0.37 | 0.36 | 0.25 | 0.73 |
| 1968 | 0.84 | 0.00 | 0.52 | 0.00 | 0.37 | 0.00 |
| 1969 | 0.74 | 0.39 | 0.37 | 0.53 | 0.24 | 0.85 |
| 1970 | 0.84 | 0.54 | 0.51 | 0.68 | 0.37 | 0.91 |
| 1971 | 0.65 | 0.63 | 0.27 | 0.75 | 0.17 | 0.94 |
| 1972 | 0.40 | 0.36 | 0.12 | 0.50 | 0.07 | 0.83 |
| 1973 | 0.58 | 0.46 | 0.22 | 0.61 | 0.14 | 0.88 |
| 1974 | 0.74 | 0.53 | 0.37 | 0.67 | 0.24 | 0.91 |
| 1975 | 0.87 | 0.49 | 0.58 | 0.63 | 0.43 | 0.90 |
| 1976 | 0.91 | 0.43 | 0.67 | 0.57 | 0.52 | 0.87 |
| 1977 | 0.67 | 0.17 | 0.29 | 0.28 | 0.19 | 0.66 |
| 1978 | 0.60 | 0.36 | 0.23 | 0.50 | 0.14 | 0.84 |
| 1980 | 0.89 | 0.30 | 0.62 | 0.43 | 0.47 | 0.79 |
| 1981 | 0.63 | 0.24 | 0.25 | 0.36 | 0.16 | 0.74 |
| 1982 | 0.88 | 0.40 | 0.60 | 0.54 | 0.45 | 0.86 |
| 1983 | 0.83 | 0.14 | 0.50 | 0.22 | 0.36 | 0.59 |
| 1984 | 0.71 | 0.15 | 0.33 | 0.24 | 0.22 | 0.61 |
| 1985 | 0.79 | 0.21 | 0.42 | 0.32 | 0.29 | 0.70 |
| 1986 | 0.67 | 0.46 | 0.29 | 0.60 | 0.18 | 0.88 |
| 1987 | 0.65 | 0.42 | 0.27 | 0.56 | 0.17 | 0.87 |
| 1988 | 0.55 | 0.44 | 0.19 | 0.58 | 0.12 | 0.87 |
| 1989 | 0.78 | 0.46 | 0.42 | 0.61 | 0.28 | 0.89 |
| 1990 | 0.77 | 0.42 | 0.41 | 0.57 | 0.28 | 0.87 |


| AVG | 0.73 | 0.36 | 0.38 | 0.48 | 0.27 | 0.78 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| AVG1 | 0.73 | 0.33 | 0.37 | 0.46 | 0.25 | 0.79 |


| Carlin Adjustment $=2.0$ <br> Evaluation of P-fraction |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: |
| Can. | Grld. | Can. | Grld. | Can. | Grld. |
| 0.1 | 0.9 | 0.5 | 0.5 | 0.9 | 0.1 |
| 0.86 | 0.38 | 0.54 | 0.53 | 0.40 | 0.85 |
| 0.91 | 0.00 | 0.68 | 0.00 | 0.54 | 0.00 |
| 0.85 | 0.56 | 0.54 | 0.69 | 0.39 | 0.92 |
| 0.91 | 0.71 | 0.68 | 0.81 | 0.54 | 0.96 |
| 0.79 | 0.77 | 0.42 | 0.86 | 0.29 | 0.97 |
| 0.57 | 0.53 | 0.21 | 0.67 | 0.13 | 0.91 |
| 0.74 | 0.63 | 0.36 | 0.75 | 0.24 | 0.94 |
| 0.85 | 0.69 | 0.54 | 0.80 | 0.39 | 0.95 |
| 0.93 | 0.66 | 0.73 | 0.78 | 0.60 | 0.95 |
| 0.95 | 0.60 | 0.80 | 0.73 | 0.69 | 0.93 |
| 0.8 | 0.30 | 0.45 | 0.43 | 0.31 | 0.79 |
| 0.75 | 0.53 | 0.37 | 0.67 | 0.25 | 0.91 |
| 0.94 | 0.46 | 0.76 | 0.61 | 0.64 | 0.88 |
| 0.77 | 0.39 | 0.41 | 0.53 | 0.28 | 0.85 |
| 0.94 | 0.57 | 0.75 | 0.70 | 0.62 | 0.92 |
| 0.91 | 0.24 | 0.67 | 0.36 | 0.53 | 0.74 |
| 0.83 | 0.26 | 0.50 | 0.38 | 0.35 | 0.76 |
| 0.88 | 0.34 | 0.60 | 0.49 | 0.45 | 0.82 |
| 0.80 | 0.63 | 0.44 | 0.75 | 0.31 | 0.94 |
| 0.79 | 0.59 | 0.42 | 0.72 | 0.29 | 0.93 |
| 0.71 | 0.61 | 0.32 | 0.74 | 0.21 | 0.93 |
| 0.88 | 0.63 | 0.59 | 0.76 | 0.44 | 0.94 |
| 0.87 | 0.60 | 0.58 | 0.73 | 0.43 | 0.93 |
|  |  |  |  |  |  |


| 0.84 | 0.51 | 0.54 | 0.63 | 0.41 | 0.86 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 0.84 | 0.49 | 0.53 | 0.62 | 0.39 | 0.88 |

AVG is average for entire time series. AVG10 is average for last 10 years.

Table 5.6.1. Catch and number of salmon sampled for fork length and cwTs during the sampling program at West Greenland, 1985-91.

| Year | Catch tonnes | Sampling Period Stweeks | \% of catch in sampling period | No. of fish sampled for fork length | \% of catch sampled for fork length | No. of fish sampled for CWTs | \% of catch sampled for CWTs |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 864 | 31-34 | 80.0 | 13,272 | 4.48 | 14,319 | 4.83 |
| 1986 | 960 | 33-35 | 92.1 | 20,394 | 6.39 | 30,360 | 9.52 |
| 1987 | 966 | 31-36 | 79.3 | 13,425 | 4.39 | 25,047 | 8.20 |
| 1988 | 893 | 33-38 | 67.1 | 11,047 | 3.93 | 22,327 | 7.95 |
| 1989 | 337 | 33-35 | 74.7 | 9,366 | 7.97 | 15,588 | 13.26 |
| 1990 | 227 | 31-33 | 25.6 | 4,897 | 5.70 | 6,410 | 7.45 |
| 1991 | 437 | 32-34 | 51.9 | 5,011 | 3.03 | 7,357 | 4.46 |

Table 5.6.2. Recoveries of Carlin tags from USA within and outside the sampling period. Binomial test of whether the reporting rate within and outside the sampling period are different.

Year No. of Carlin No. of Carlin \% of Carlin \% of catch Prob. of no diff. tags in tags outside tags within within in reporting rate sampling $p$. sampling $p$. sampling p. sampling $p$. within and
outside sampling p .

| 1985 | 27 | 6 | 81.8 | 80.0 | 0.79 |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | 55 | 10 | 84.6 | 92.1 | 0.03 |
| 1987 | 112 | 33 | 79.3 | 77.2 | 0.54 |
| 1988 | 44 | 41 | 51.8 | 67.1 | $<0.01$ |
| 1989 | 59 | 17 | 77.6 | 74.7 | 0.56 |
| 1990 | 3 | 11 | 21.4 | 25.6 | 0.99 |
| 1991 | 5 | 5 | 50.0 | 51.9 | 0.90 |

Table 6.1.1. Nominal catch in tonnes of Atlantic salmon of all ages for Salmon Fishing Areas of Newfoundland and Labrador and Quebec commercial fisheries in 1986-91. Numbers in parentheses are catch totals in numbers of fish. Catches for 1991 are preliminary.

| Salmon Fishing Areas | 1986 | 1987 | 1988 | 1989 | 1990 |  | 1991 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | Catch | Quota | Catch | Quota |
| 1 | 89 | 75 | 65 | 76 | 30 | $80^{\text {a }}$ | 7 | $80^{\text {a }}$ |
| 2 | 308 | 407 | 292 | 213 | 149 | 200 | 79 | 200 |
| 3 | 192 | 369 | 192 | 151 | 136 | 155 | 108 | 120 |
| 4 | 200 | 180 | 104 | 133 | 93 | 100 | 52 | 78 |
| 5 | 61 | 60 | 39 | 37 | 24 | 25 | 18 | 25 |
| 6 | 54 | 48 | 25 | 27 | 17 | 20 | 19 | 20 |
| 7-11 | 167 | 137 | 82 | 108 | 82 | 82 | 70 | 82 |
| 13-14 | 159 | 212 | 174 | 122 | 87 | 95 | 81e | 75 |
| Q7-Q9 | 85 | 97 | 89 | 79 | 64 |  | 77 |  |
|  | $(21,802) \mathrm{c}$ | $(23,525) \mathrm{c}$ | $(22,863) \mathrm{c}$ | $(20,525) \mathrm{c}$ | $(19,272)$ | $(30,060) \mathrm{b}$ | $(21,601)$ | $(28,359) \mathrm{b}$ |
| Q11 | $\begin{aligned} & 15 d \\ & (2,794) \end{aligned}$ | $\begin{gathered} 11 \mathrm{~d} \\ (2,212) \end{gathered}$ | $\begin{gathered} 9 \text { 9d } \\ (1,647) \end{gathered}$ | $\begin{array}{r} 1 d \\ (245) \end{array}$ | $\begin{array}{r} 1 \\ (225) \end{array}$ | (30, 15 | $\begin{array}{r} 1 \\ (277) \end{array}$ | (28, 15 |
| Total | 1329 | 1596 | 1071 | 947 | 683 |  | 512 |  |

a Allowance.
b Quota was in numbers.
C Quotas for 1986 to 1989 were 33,125 per year.
d Quota was 15 t each year.
e 17 tonnes of SFA 14 catch was monitored as part of SFA 2 quota.

Table 6.1.2. Preliminary catches of Atlantic salmon in Canada in 1991 (in kg round fresh weight).

|  | Small Salmon |  | Large Salmon |  | $\frac{\text { Total }}{(\mathrm{kg})}$ | $\frac{\text { Total }}{(\%)}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | (kg) | (\%) | (kg) | (\%) |  |  |
| QUEBEC |  |  |  |  |  |  |
| R | 11,140 | 3.5 | 60,363 | 16.7 | 71,503 | 10.5 |
| N | 152 | 0.0 | 15,983 | 4.4 | 16,135 | 2.4 |
| C | 6,210 | 2.0 | 72,190 | 20.0 | 78,400 | 11.5 |
| Total | 17,502 | 5.5 | 148,536 | 41.1 | 166,038 | 24.5 |
| NFLD |  |  |  |  |  |  |
| R | 28,637 | 9.0 | 300 | 0.1 | 28,937 | 4.3 |
| N | 486 | 0.2 | 0 | 0.0 | 486 | 0.1 |
| C | 229,992 | 72.4 | 203,128 | 56.2 | 433,120 | 63.8 |
| Total | 259,115 | 81.6 | 203,428 | 56.3 | 462,543 | 68.1 |
| NB |  |  |  |  |  |  |
| R | 31,637 | 10.0 | 0 | 0.0 | 31,637 | 4.7 |
| N | 2,469 | 0.8 | 7,701 | 2.1 | 10,170 | 1.5 |
| C | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 |
| Total | 34,106 | 10.8 | 7,701 | 2.1 | 41,807 | 6.2 |
| NS |  |  |  |  |  |  |
| R | 6,195 | 2.0 | 0 | 0.0 | 6,195 | 0.9 |
| N | 774 | 0.2 | 1,527 | 0.4 | 2,301 | 0.3 |
| C | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 |
| Total | 6,969 | 2.2 | 1,527 | 0.4 | 8,496 | 1.2 |
| PEI |  |  |  |  |  |  |
| R* | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 |
| N | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 |
| C | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 |
| Total | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 |
| TOTAL |  |  |  |  |  |  |
| R | 77,609 | 24.4 | 60,663 | 16.8 | 138,272 | 20.4 |
| N | 3,881 | 1.2 | 25,211 | 7.0 | 29,092 | 4.3 |
| C | 236,202 | 74.3 | 275,318 | 76.2 | 511,520 | 75.3 |
| Total | 317,692 | 100.0 | 361,192 | 100.0 | 678,884 | 100.0 |

$R=$ Recreational
$\mathrm{N}=$ Native Food
C = Commercial

* No catch statistics collected.

Table 6.1.3. Nominal catches (tonnes) in Newfoundland and Labrador commercial Atlantic salmon fishery, 1971-91. Catches for 1991 are preliminary.

| Year | Small | Large | Total ${ }^{\text {a }}$ | Quota ${ }^{\text {b }}$ |
| :---: | :---: | :---: | :---: | :---: |
| 1971 | - | - | 1577 | - |
| 1972 | - | - | 1394 | - |
| 1973 | - | - | 2011 | - |
| 1974 | - | - | 2010 | - |
| 1975 | 750 | 1294 | 2043 | - |
| 1976 | 632 | 1380 | 2013 | - |
| 1977 | 533 | 1404 | 1938 | - |
| 1978 | 274 | 907 | 1180 | - |
| 1979 | 494 | 495 | 987 | _ |
| 1980 | 809 | 1295 | 2103 | _ |
| 1981 | 676 | 1233 | 1910 | - |
| 1982 | 578 | 743 | 1321 | - |
| 1983 | 417 | 611 | 1028 | - |
| 1984 | 332 | 465 | 797 | _ |
| 1985 | 470 | 411 | 881 | _ |
| 1986 | 608 | 622 | 1230 | - |
| 1987 | 705 | 780 | 1485 | - |
| 1988 | 511 | 461 | 972 | - |
| 1989 | 431 | 436 | 867 | - |
| 1990 | 284 | 334 | 618 | 677 |
| 1991 | 230 | 203 | 434 | 600 |

a Differences between total and sum of small and large are due to rounding.
${ }^{\mathrm{b}}$ Excludes an allowance of 80 t for SFA 1.

Table 6.1.4. Preliminary landings, in tonnes, of Atlantic salmon harvested in the commercial fisheries in Atlantic Canada, 1991. Comparisons are made to the average landings, 1984-89.

| SFA | Small |  | Large |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Weight(t) | $\begin{array}{r} \% \text { Change } \\ 1984-89 \end{array}$ | Weight(t) | $\begin{array}{r} \% \text { Change } \\ 1984-89 \end{array}$ | Weight(t) | $\begin{array}{r} \circ \text { Change } \\ 1984-89 \end{array}$ |
| 1 | 2 | -90 | 5 | -90 | 7 | -90 |
| 2 | 41 | -53 | 38 | -77 | 79 | -69 |
| 3 | 52 | -48 | 56 | -41 | 108 | -45 |
| 4 | 26 | -70 | 25 | -54 | 52 | -63 |
| 5 | 12 | -62 | 6 | -74 | 18 | -67 |
| 6 | 12 | -54 | 6 | -63 | 19 | -55 |
| 7 | 6 | -12 | 6 | -53 | 12 | -39 |
| 8 | 3 | -74 | 4 | -66 | 7 | -70 |
| 9 | 5 | -25 | 1 | -50 | 5 | -44 |
| 10 | 12 | -52 | 6 | -40 | 18 | -49 |
| 11 | 8 | -70 | 20 | -21 | 28 | -46 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 25 | -35 | 4 | -80 | 29 | -50 |
| 14 | 26 | -39 | 26 | -37 | 52 | -38 |
| 15-23 | 0 | 0 | 0 | 0 | 0 | 0 |
| Q1-6 | 0 | 0 | 0 | 0 | 0 | 0 |
| Q7 | <1 | -851 | 7 | -251 | 7 | $-30^{1}$ |
| Q8 | <1 | -641 | 29 | -191 | 29 | -211 |
| Q9 | 6 | 41 | 35 | $-2^{1}$ | 41 | $-1^{1}$ |
| Q10 | 0 | 0 | 0 | 0 | 0 | 0 |
| Q11 | 0 | - | 1 | - | 1 | $-81^{1}$ |

[^4]Table 6.1.5. Preliminary landings, in numbers, of Atlantic salmon harvested in the recreational fisheries in Atlantic Canada, 1991. Comparisons are made to the average landings, 1986-90.

| SFA | Small |  | Large |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Numbers | $\begin{array}{r} \text { \% Change } \\ 1986-90 \end{array}$ | Numbers | $\begin{array}{r} \% \text { Change } \\ 1986-90 \end{array}$ | Numbers | $\begin{array}{r} \% \text { Change } \\ 1986-90 \end{array}$ |
| 1 | 29 | -961 | 1 | -991 | 30 | -961 |
| 2 | 1,122 | -481 | 27 | -871 | 1,149 | -521 |
| 3 | 1,316 | 181 | 0 | 0 | 1,316 | 181 |
| 4 | 4,892 | -461 | 0 | 0 | 4,892 | -461 |
| 5 | 2,048 | -351 | 0 | 0 | 2,048 | -351 |
| 6 | 186 | -501 | 0 | 0 | 186 | -501 |
| 7 | 36 | -641 | 0 | 0 | 36 | -64 1 |
| 8 | 11 | -891 | 0 | 0 | 11 | -891 |
| 9 | 560 | -691 | 0 | 0 | 560 | -691 |
| 10 | 230 | -821 | 0 | 0 | 230 | -82 ${ }^{1}$ |
| 11 | 1,853 | -65 ${ }^{1}$ | 0 | 0 | 1,853 | -65 ${ }^{1}$ |
| 12 | 644 | -30 | 0 | 0 | 644 | -30 |
| 13 | 5,188 | -17 | 0 | 0 | 5,188 | -17 |
| 14 | 4,717 | -24 | 49 | -69 | 4,766 | -24 |
| 15 | 2,894 | -46 | 0 | 0 | 2,899 | -46 |
| 16 | 10,049 | -39 | 0 | 0 | 10,049 | -39 |
| $17^{3}$ | - | - | 0 | 0 | 10,0 | - |
| 18 | 1,230 | 3 | 0 | 0 | 1,230 | 3 |
| 19 | 446 | -46 | 0 | 0 | , 446 | -46 |
| 20 | 1,414 | -40 | 0 | 0 | 1,414 | -40 |
| 21 | 558 | -85 | 0 | 0 | 558 | -85 |
| 22 | 10 | N/A ${ }^{2}$ | 0 | 0 | 10 | $\mathrm{N} / \mathrm{A}^{2}$ |
| 23 | 1,945 | -43 | 0 | 0 | 1,945 | -43 |
| Q1 | 1,124 | -2 | 2,555 | -15 | 3,679 | -11 |
| Q2 | 435 | 16 | 1,605 | 16 | 2,040 | 16 |
| Q3 | 795 | 11 | 1,298 | 23 | 2,093 | 18 |
| Q4 | 0 | - | 0 | - | 0 | - |
| Q5 | 55 | 62 | 108 | 184 | 163 | 126 |
| Q6 | 364 | 29 | 465 | -16 | 829 | -1 |
| Q7 | 631 | -42 | 476 | -22 | 1,107 | -35 |
| Q8 | 755 | -14 | 3,587 | 35 | 4,342 | 23 |
| Q9 | 1,152 | 0 | 109 | -81 | 1,261 | -27 |
| Q10 | 244 | -47 | 559 | -14 | 803 | -27 |
| Q11 | 576 | -47 | 380 | -54 | 947 | -50 |

[^5]Table 6.1.1.1. The number of salmon examined for CWTs, periods of sampling and origin of tags recovered in Canada, 1991.

| Location | Sampling <br> Period | Number <br> Examined | Adipose <br> Clipped | Percent <br> Clipped | Origin | Total <br> CWanada | Percent <br> CWT |  |
| :--- | :--- | ---: | ---: | :--- | :--- | :--- | :--- | :--- |
| Campbellton | $6 / 25-8 / 7$ | 294 | 2 | 0.68 | 0 | 0 | 0 | 0.00 |
| Conche | $6 / 29-8 / 12$ | 3653 | 44 | 1.20 | 1 | 40 | 41 | 1.12 |
| Croque | $7 / 19-8 / 6$ | 2256 | 25 | 1.11 | 1 | 20 | 21 | 0.93 |
| Englee | $6 / 17-8 / 2$ | 1472 | 5 | 0.34 | 0 | 1 | 1 | 0.07 |
| Fogo | $7 / 1-8 / 8$ | 425 | 4 | 0.94 | 0 | 0 | 0 | 0.00 |
| Fox Harbour | $7 / 16-8 / 7$ | 1114 | 3 | 0.27 | 0 | 0 | 0 | 0.00 |
| Goose Cove | $7 / 18-7 / 20$ | 6458 | 4 | 0.06 | 0 | 4 | 4 | 0.06 |
| Harbour Deep | $7 / 8-7 / 31$ | 1084 | 38 | 3.51 | 2 | 25 | 27 | 2.49 |
| Leading Tickles | $7 / 2-7 / 17$ | 65 | 7 | 10.77 | 0 | 1 | 1 | 1.54 |
| Musgrave | $7 / 12-7 / 17$ | 33 | 0 | 0.00 | 0 | 0 | 0 | 0.00 |
| Shoe Cove | $7 / 4-7 / 29$ | 1529 | 21 | 1.37 | 0 | 19 | 19 | 1.24 |
| Twillingate | $6 / 23-7 / 19$ | 778 | 6 | 0.77 | 0 | 6 | 6 | 0.77 |
| Total |  | 19161 | 159 |  | 4 | 116 | 120 |  |
| Average |  |  | 1.75 |  |  |  | 0.69 |  |

Table 6.1.2.1. Carlin tag returns from 1SW salmon of Maine origin in Newfoundland and Labrador by year and Salmon Fishing Area. ( $99=$ unknown area)

|  | SFA |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 99 | TOT |
| 1967 | 3 | 1 | 7 | 14 | 5 | 0 | 4 | 0 | 1 | 1 | 2 | 0 | 0 | 0 | 2 | 40 |
| 1968 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 1969 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 1970 | 5 | 2 | 13 | 5 | 1 | 1 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 29 |
| 1971 | 10 | 2 | 4 | 18 | 10 | 3 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 48 |
| 1972 | 6 | 1 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 12 |
| 1973 | 6 | 1 | 6 | 4 | 1 | 1 | 1 | 3 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 25 |
| 1974 | 0 | 5 | 19 | 38 | 13 | 10 | 5 | 3 | 3 | 3 | 0 | 1 | 0 | 3 | 0 | 103 |
| 1975 | 16 | 4 | 18 | 36 | 13 | 6 | 1 | 4 | 1 | 2 | 0 | 0 | 0 | 1 | 0 | 102 |
| 1976 | 18 | 6 | 26 | 14 | 5 | 5 | 0 | 0 | 0 | 3 | 2 | 0 | 0 | 1 | 0 | 80 |
| 1977 | 2 | 1 | 6 | 5 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 16 |
| 1978 | 4 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 |
| 1980 | 55 | 24 | 112 | 72 | 22 | 6 | 0 | 3 | 2 | 3 | 12 | 0 | 0 | 3 | 1 | 315 |
| 1981 | 14 | 0 | 2 | 10 | 7 | 5 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 41 |
| 1982 | 14 | 7 | 20 | 21 | 7 | 6 | 1 | 0 | 0 | 1 | 4 | 0 | 2 | 2 | 0 | 85 |
| 1983 | 8 | 1 | 11 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 26 |
| 1984 | 12 | 4 | 7 | 7 | 4 | 2 | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 39 |
| 1985 | 20 | 3 | 15 | 36 | 11 | 1 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 94 |
| 1986 | 3 | 5 | 6 | 2 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 21 |
| 1987 | 14 | 2 | 16 | 4 | 6 | 2 | 0 | 2 | 1 | 1 | 0 | 0 | 0 | 0 | 1 | 49 |
| 1988 | 8 | 2 | 5 | 0 | 1 | 0 | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 20 |
| 1989 | 25 | 5 | 10 | 6 | 4 | 1 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 54 |
| 1990 | 0 | 2 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 8 |
| 1991 | 0 | 0 | 12 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14 |
| unk | 2 | 0 | 1 | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 8 |
| TOT | 245 | 78 | 321 | 304 | 117 | 49 | 20 | 22 | 10 | 19 | 27 | 2 | 2 | 19 | 5 | 1240 |

Table 6.1.2.2. Estimated harvest of 1SW salmon of Maine origin in Newfoundland and Labrador by year and Salmon Fishing Area. ( $99=$ unknown area)

| SFA |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| YEAR | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 99 |  |
| 1967 | 14 | 5 | 43 | 87 | 31 | 0 | 25 | 0 | 6 | 6 | 12 | 0 | 0 | 0 | 12 | 242 |
| 1968 | 0 | 0 | 0 | 274 | 137 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 411 |
| 1969 | 0 | 0 | 185 | 0 | 0 | 0 | 0 | 92 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 277 |
| 1970 | 56 | 23 | 188 | 72 | 14 | 14 | 0 | 14 | 0 | 14 | 0 | 0 | 0 | 0 | 0 | 398 |
| 1971 | 51 | 10 | 26 | 117 | 65 | 20 | 0 | 0 | 0 | 0 | 7 | 0 | 0 | 0 | 0 | 295 |
| 1972 | 47 | 8 | 0 | 0 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 10 | 0 | 0 | 0 | 105 |
| 1973 | 44 | 7 | 56 | 38 | 9 | 9 | 9 | 28 | 9 | 0 | 0 | 0 | 0 | 9 | 0 | 220 |
| 1974 | 0 | 29 | 141 | 283 | 97 | 74 | 37 | 22 | 22 | 22 | 0 | 7 | 0 | 22 | 0 | 758 |
| 1975 | 129 | 32 | 187 | 374 | 135 | 62 | 10 | 42 | 10 | 21 | 0 | 0 | 0 | 10 | 0 | 1014 |
| 1976 | 418 | 139 | 777 | 418 | 149 | 149 | 0 | 0 | 0 | 90 | 60 | 0 | 0 | 30 | 0 | 2230 |
| 1977 | 95 | 48 | 368 | 307 | 0 | 0 | 0 | 0 | 0 | 61 | 61 | 0 | 0 | 0 | 0 | 940 |
| 1978 | 234 | 0 | 75 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 309 |
| 1980 | 666 | 291 | 1744 | 1121 | 343 | 93 | 0 | 47 | 31 | 47 | 187 | 0 | 0 | 47 | 16 | 4631 |
| 1981 | 330 | 0 | 61 | 303 | 212 | 151 | 30 | 0 | 30 | 0 | 0 | 0 | 0 | 30 | 0 | 1147 |
| 1982 | 217 | 109 | 399 | 419 | 140 | 120 | 20 | 0 | 0 | 20 | 80 | 0 | 40 | 40 | 0 | 1603 |
| 1983 | 441 | 55 | 779 | 425 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1700 |
| 1984 | 350 | 117 | 262 | 262 | 150 | 75 | 37 | 0 | 0 | 37 | 37 | 0 | 0 | 0 | 0 | 1329 |
| 1985 | 400 | 60 | 386 | 926 | 283 | 26 | 77 | 51 | 0 | 0 | 0 | 0 | 0 | 51 | 26 | 2288 |
| 1986 | 67 | 112 | 172 | 57 | 29 | 0 | 0 | 29 | 0 | 0 | 0 | 0 | 0 | 86 | 0 | 552 |
| 1987 | 139 | 20 | 204 | 51 | 77 | 26 | 0 | 26 | 13 | 13 | 0 | 0 | 0 | 0 | 13 | 580 |
| 1988 | 138 | 34 | 111 | 0 | 22 | 0 | 22 | 0 | 0 | 22 | 22 | 0 | 0 | 22 | 0 | 393 |
| 1989 | 708 | 142 | 364 | 218 | 146 | 36 | 36 | 36 | 0 | 36 | 0 | 0 | 0 | 0 | 0 | 1722 |
| 1990 | 0 | 161 | 207 | 207 | 0 | 0 | 0 | 0 | 0 | 0 | 207 | 0 | 0 | 0 | 0 | 780 |
| TOT | 4544 | 1400 | 6736 | 5960 | 2078 | 857 | 305 | 388 | 123 | 390 | 673 | 17 | 40 | 349 | 66 | 23926 |

Table 6.1.2.3. Estimated total run size of 1 SW and 2 SW salmon returning to Maine rivers and estimated harvests of 1SW salmon in Newfoundland and Labrador fisheries. All run size and harvest estimates are computed assuming 85 percent fish passage efficiency.

| Year$\qquad$ | Run |  |  | Harvest <br> i | Harves /2SW Run Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & \hline \text { 1SW } \\ & i \\ & \hline \end{aligned}$ | $\begin{aligned} & 2 S W \\ & i+1 \end{aligned}$ | 1/2SW <br> Ratio |  |  |
| 1967 | 100 | 729 | 0.138 | 242 | 0.332 |
| 1968 | 24 | 690 | 0.035 | 411 | 0.595 |
| 1969 | 36 | 856 | 0.041 | 277 | 0.324 |
| 1970 | 14 | 687 | 0.021 | 398 | 0.579 |
| 1971 | 44 | 1449 | 0.030 | 295 | 0.204 |
| 1972 | 32 | 1448 | 0.022 | 105 | 0.072 |
| 1973 | 43 | 1411 | 0.030 | 220 | 0.156 |
| 1974 | 99 | 2345 | 0.042 | 758 | 0.323 |
| 1975 | 116 | 1341 | 0.086 | 1014 | 0.756 |
| 1976 | 231 | 2025 | 0.114 | 2230 | 1.101 |
| 1977 | 98 | 4145 | 0.024 | 940 | 0.227 |
| 1978 | 161 | 1878 | 0.086 | 309 | 0.165 |
| 1979 | 251 | 5662 | 0.044 | NA | NA |
| 1980 | 847 | 5122 | 0.165 | 4631 | 0.904 |
| 1981 | 1148 | 6023 | 0.191 | 1147 | 0.191 |
| 1982 | 315 | 1930 | 0.163 | 1603 | 0.831 |
| 1983 | 271 | 3045 | 0.089 | 1700 | 0.558 |
| 1984 | 388 | 4855 | 0.080 | 1329 | 0.274 |
| 1985 | 337 | 5568 | 0.061 | 2288 | 0.411 |
| 1986 | 711 | 2397 | 0.297 | 552 | 0.230 |
| 1987 | 950 | 2855 | 0.333 | 580 | 0.203 |
| 1988 | 896 | 2946 | 0.304 | 393 | 0.134 |
| 1989 | 1267 | 4370 | 0.290 | 1722 | 0.394 |
| 1990 | 654 | 2057 | 0.318 | 780 | 0.379 |

$N A=$ Not Available since no smolts were tagged in 1978.

Table 6.1.2.4. CWT and Carlin tag based harvest estimates for sampled areas in Canada, 1990.

|  |  |  | HARVEST |  | CWT | METHODS |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | LOCATION (1) | CATCH | SCAN | CWT | CARLIN | CV |
|  | RATIO |  |  |  |  |  |

COMMUNITIES

| GOOSE COVE (SS 3) | 4710 | 5135 | 7.8 | 0.0 | 56 |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| CROQUE (SS 3) | 2253 | 1752 | 14.8 | 0.0 |  |  |
| CONCHE (SS3) | 5876 | 1567 | 0.0 | 0.0 |  |  |
| ENGLEE (SS 4) | 2575 | 1936 | 22.9 | 0.0 | 24 |  |
| HARBOUR DEEP (SS 4) | 669 | 545 | 37.8 | 0.0 | 12 |  |
| SHOE COVE (SS 6) | 5214 | 4746 | 36.7 | 103.0 |  | 0.36 |
| LEADING TICKLES (SS 6) | 708 | 593 | 15.3 | 0.0 |  |  |
| TWILLINGATE (SS 7) | 1093 | 1112 | 0.0 | 0.0 |  |  |
| FOGO (SS 8) | 503 | 864 | 2.7 | 0.0 |  |  |
| MUSGRAVE HARBOUR (SS 9) | 88 | 99 | 0.0 | 0.0 |  |  |
| FOX HARBOUR (SS 52) | 856 | 1566 | 0.0 | 0.0 |  |  |

STATISTICAL SECTIONS

| SS 3 | 10556 | 6887 | 33.3 | 0.0 | 18 |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| SS 4 | 18329 | 4048 | 137.2 | 103.0 | 7 | 1.33 |
| SS 6 | 16086 | 5339 | 173.8 | 103.3 | 9 | 1.68 |
| SS 6 + UNK |  |  | 173.8 | 206.6 | 9 | 0.84 |
| SS 7 | 16086 | 1112 | 0.0 | 0.0 |  |  |
| SS 8 | 2953 | 864 | 16.1 | 0.0 | 22 |  |
| SS 9 | 4573 | 99 | 0.0 | 0.0 |  |  |
| SS 52 | 9708 | 1566 | 0.0 | 80.3 |  |  |

SALMON FISHING AREAS

| SFA 3 | 40408 | 10935 | 200.4 | 103.3 | 6 | 1.94 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| SFA 4 | 35615 | 7414 | 378.6 | 103.3 | 6 | 3.66 |
| SFA 4 + UNK |  |  | 378.6 | 206.6 | 6 | 1.83 |

(1) UNK indicates Carlin estimates include tags of unknown recovery date.

Table 6.1.3.1. Exploitation rates from the recreational fisheries for Atlantic salmon from several Gulf Region rivers. Unadjusted and adjusted rates are based on mark/recapture techniques. Adjusted rates account for reporting rate and tag loss estimates. Exploitation rate on large salmom are for hook and release fish. N/A = insufficient data for estimation.

| River | Size | Year | Number Tagged | Unadjusted Rate |  |  | Adjusted Rate | Angling/ <br> Returns |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Early | Late | Combined |  |  |
| $\begin{aligned} & \text { Miramichi } \\ & (\text { SFA 16) } \end{aligned}$ | Small | 1986 | 1587 | 0.10 | 0.05 | 0.09 | 0.20 | 0.22 |
|  | ( $<63 \mathrm{~cm}$ ) | 1987 | 1103 | 0.09 | 0.02 | 0.08 | 0.17 | 0.24 |
|  |  | 1988 | 1278 | 0.15 | 0.14 | 0.15 | 0.30 | 0.25 |
|  |  | 1989 | 833 | 0.10 | 0.05 | 0.10 | 0.20 | 0.32 |
|  |  | 1990 | 693 | 0.12 | 0.09 | 0.17 | 0.38 | 0.26 |
|  |  | 1991 | 559 | 0.10 | 0.04 | 0.08 | 0.16 | 0.19 |
|  | Large | 1986 | 400 | 0.04 | 0.05 | 0.04 | N/A | 0.46 |
|  |  | 1987 | 275 | 0.01 | 0.02 | 0.01 | N/A | 0.62 |
|  |  | 1988 | 241 | 0.02 | 0.03 | 0.03 | N/A | 0.47 |
|  |  | 1989 | 206 | 0.03 | 0.02 | 0.03 | N/A | 0.70 |
|  |  | 1990 | 338 | 0.04 | 0.05 | 0.05 | N/A | 0.33 |
|  |  | 1991 | 317 | 0.04 | 0.02 | 0.03 | N/A | 0.21 |
| Margaree (SFA 18) | Small | 1988 | 173 |  | 0.08 |  | 0.16 |  |
|  | (<63cm) | 1989 | 78 |  | 0.05 |  | 0.10 |  |
|  |  | 1990 | 120 |  | 0.11 |  | 0.19 |  |
|  |  | 1991 | 164 | 0.15 |  |  | 0.19 |  |
|  | Large | 1988 | 155 |  | 0.04 |  | 0.08 |  |
|  |  | 1989 | 347 |  | 0.04 |  | 0.07 |  |
|  |  | 1990 | 217 |  | 0.08 |  | 0.13 |  |
|  |  | 1991 | 203 | 0.12 |  |  | 0.17 |  |
| Humber <br> (SFA 13) | Small | 1990 | 214 | 0.13 |  |  | 0.25 |  |
|  | ( $<63 \mathrm{~cm}$ ) | 1991 | 64 | 0.16 |  |  | N/A |  |

Table 6.4.1. Quotas, closing date and nominal landings for the commercial fisheries, SFA, 1-14 in 1990 and 1991. The commercial fisheries opened June 5 in all SFAs. The landings for 1991 are preliminary.

| SFA | 1990 |  |  | 1991 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Quota ( $t$ ) | Date Closed | $\begin{aligned} & \text { Landings } \\ & (t) \end{aligned}$ | Quota ( t ) | Date Closed | $\begin{gathered} \text { Landings } \\ \hline() \end{gathered}$ |
| 1 | a | Oct. 15 | 30 | a | Oct 15 | 7 |
| 2 | 200 | Oct. 15 | 149 | 200 | Oct 15 | 79 |
| 3 | 155 | Oct. 15 | 136 | 120 | Sep 23 | 108 |
| 4 | 100 | July 25 | 93 | 78 | Oct 15 | 52 |
| 5 | 25 | July 7 | 24 | 25 | Jul 25 | 18 |
| 6 | 20 | July 14 | 17 | 20 | Jul 21 | 19 |
| 7 | 15 | Oct. 15 | 12 | 15 | Jul 27 | 12 |
| 8 | 10 | July 21 | 9 | 10 | Jul 26 | 7 |
| 9 | 7 | Oct. 15 | 7 | 7 | Jul 13 | 5 |
| 10 | 25 | June 30 | 24 | 25 | Jul 11 | 18 |
| 11 | 25 | June 21 | 30 | 25 | Jun 29 | 28 |
| 12 | 0 | not open | 0 | 0 | not open | 0 |
| 13 | 35 | July 3 | 43 | 25 | Jul 6 | 29 |
| 14 | 60 | July 14b | 44 | 50 | Oct 15 | 52 |
| Total | 677 |  | 618 c | 600 |  | 434 c |

[^6]Table 6.4.1.1. Estimates of small and large salmon by weight which were not harvested in (a) 1990 and (b) 1991 due to the early closure of the commercial fisheries. Number in parenthesis is the percent of the harvest which would not have been caught, 1984-89 if the season was closed on the same date as in 1990 in (a) and 1991 in (b).

(a) 1990

Small Salmon

| 4 | 62 | 5 | 20 | 10 | (14) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 15 | 4 | 17 | 9 | (38) |
| 6 | 10 | 2 | 10 | 5 | (31) |
| 8 | 4 | <1 | <1 | <1 | (11) |
| 10 | 18 | 10 | 39 | 16 | (47) |
| 11 | 9 | 5 | 71 | 14 | (60) |
| 13 | 29 | 8 | 30 | 13 | (31) |
| 14 | 25 | 4 | 46 | 11 | (31) |
| Total | 172 |  |  | 79 |  |


| Large Salmon |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: |
| 4 | 31 | 5 | 10 | $6(17)$ |
| 5 | 10 | 1 | 5 | $3(24)$ |
| 6 | 7 | $<1$ | 2 | $(21)$ |
| 8 | 5 | 1 | 1 | $(5)$ |
| 10 | 6 | 3 | 6 | $3(31)$ |
| 11 | 14 | 5 | 8 | $13(39)$ |
| 13 | 19 | 3 | 16 | $4(22)$ |
| 14 | 113 |  |  | $\frac{7}{39}(27)$ |
| Total |  |  |  |  |

(b) 1991

| Small Salmon 52 ( |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |
| 5 | 12 | <1 | 3 | 2 | (11) |
| 6 | 12 | 1 | 7 | 3 | (20) |
| 7 | 6 | <1 | <1 | <1 | (5) |
| 8 | 3 | <1 | <1 | <1 | (6) |
| 9 | 5 | 1 | 5 | 2 | (32) |
| 10 | 12 | 3 | 5 | 4 | (24) |
| 11 | 8 | 2 | 11 | 4 | (36) |
| 13 | 25 | 4 | 17 | 6 | (21) |
| Total | 135 |  |  | 21 |  |


| Large Salmon |  |
| :---: | ---: |
| 3 | 56 |
| 5 | 6 |
| 6 | 6 |
| 7 | 6 |
| 8 | 4 |
| 9 | 1 |
| 10 | 6 |
| 11 | 20 |
| 13 | $\underline{4}$ |
| Total | 109 |


| $<1$ | 3 | $<1$ | $(1)$ |
| ---: | ---: | ---: | :--- |
| $<1$ | $<1$ | $<1$ | $(9)$ |
| $<1$ | 3 | 1 | $(15)$ |
| $<1$ | 1 | $<1$ | $(7)$ |
| $<1$ | $<1$ | $<1$ | $(3)$ |
| $<1$ | 1 | $<1$ | $(26)$ |
| $<1$ | 3 | 1 | $(14)$ |
| 1 | 12 | 5 | $(19)$ |
| $<1$ | 2 | $\frac{<1}{9}(17)$ |  |

Table 6.4.1.2 Estimates of numbers of small and large salmon which were not harvested in (a) 1990 and (b) 1991 due to the early closure of the commercial fisheries. Number in parenthesis is the percent of the harvest which would not have been caught, 1984-89 if the season was closed on the same date as in 1990 in (a) and 1991 in (b).

| SFA | Harvest (Numbers) | Predicted Numbers not Caught |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Min | Max | Mean |  |
| (a) 1990 |  |  |  |  |  |
| Small Salmon |  |  |  |  |  |
| 4 | 32,648 | 2,647 | 10,508 | 5,429 | (14) |
| 5 | 8,018 | 2,321 | 9,285 | 4,932 | (38) |
| 6 | 5,187 | 1,075 | 5,135 | 2,383 | (31) |
| 8 | 2,525 | 201 | 455 | 295 | (11) |
| 10 | 10,283 | 5,641 | 22,727 | 9,225 | (47) |
| 11 | 4,845 | 2,567 | 38,106 | 7,310 | (60) |
| 13 | 16,638 | 4,411 | 16,490 | 6,944 | (29) |
| 14 | 12,216 | 2,086 | 20,000 | 5,035 |  |
| Total | 92,358 |  |  | 41,553 |  |
| Large Salmon |  |  |  |  |  |
| 4 | 6,940 | 1,112 | 2,226 | 1,408 | (17) |
| 5 | 2,603 | 361 | 1,236 |  | (24) |
| 6 | 1,534 | 182 | 1,364 |  | (21) |
| 8 | 1,100 | 20 | 175 | 57 | (5) |
| 10 | 1,825 | 390 | 1,832 | 805 | (31) |
| 11 | 4,354 | 708 | 10,686 | 2,742 | (39) |
| 13 | 3,239 | 423 | 1,682 | 988 | (23) |
| 14 | 3,608 | 601 | 2,997 | 1,338 | (27) |
| Total | 25,203 |  |  | 8,615 |  |
| (b) 1991 |  |  |  |  |  |
| Small Salmon |  |  |  |  |  |
| 3 | 27,722 | 11 | 78 |  | (<1) |
| 5 | 6,781 | 400 | 1,513 |  | (11) |
| 6 | 6,926 | 643 | 3,783 | 1,749 | (20) |
| 7 | 3,187 | 71 | 266 | 172 | (5) |
| 8 | 1,904 | 94 | 194 | 122 | (6) |
| 9 | 3,080 | 504 | 3,202 | 1,425 | (32) |
| 10 | 6,533 | 1,515 | 2,921 | 2,071 | (24) |
| 11 | 4,158 | 1,057 | 5,750 | 2,276 | (35) |
| 13 | 15,688 | 2,274 | 9,960 | 3,895 | (20) |
| Total | 75,979 |  |  | 12,572 |  |
| Large Salmon |  |  |  |  |  |
| 3 | 13,696 | 43 | 772 |  |  |
| 5 | 2,066 | 124 | 320 | 205 |  |
| 6 | 1,726 | 130 | 927 | 316 | (15) |
| 7 | 1,534 | 40 | 301 | 109 | (7) |
| 8 | 959 | 8 | 64 | 25 | (3) |
| 9 | 176 | 6 | 227 |  | (26) |
| 10 | 1,859 | 134 | 775 | 300 | (14) |
| 11 | 4,616 | 279 | 2,669 | 1,098 | (19) |
| 13 | -871 | 77 | 321 | 168 | (16) |
| Total | 27,503 |  |  | 2,471 |  |

Table 6.4.2.1. Estimates of numbers of 1 SW salmon of Maine origin which would not have been harvested in the Newfoundland-Labrador commercial fisheries, 1984-89, if the fisheries were closed each year on the same day that the fisheries were closed in 1990. The number in parenthesis is the percentage of the total numbers of 1 SW Maine-origin salmon caught in SFA.

| SFA | Date f closed |  |  |  |  |  | Year |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1990 | 198 |  | 1985 |  | 1986 |  | 1987 |  | 1988 |  | 1989 |  | Mean | (\%) |
| 4 | Jul 25 |  | (50) | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  |  | (8) |
| 5 | Jul 07 | 111 | (100) |  | (100) | 0 |  | 4 | (6) | 0 |  | 36 | (33) | 38 | (58) |
| 6 | Jul 14 |  | (50) |  | (100) | 0 |  | 2 | (7) | 0 |  | 0 |  | 11 | (52) |
| 8 | Jul 24 | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  |
| 10 | Jun 30 | 5 | (14) | 0 |  | 0 |  | 13 | (100) | 0 |  | 36 | (100) | 9 | (50) |
| 11 | Jun 21 | 37 | (100) | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 6 (10 | (100) |
| 13 | Jul 03 | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  | 0 |  |
| 14 | Jul 14 | 0 |  |  | (100) | 86 | (100) | 0 |  | 22 | (100) | 0 |  | 27 (100 | (100) |
| Total |  | 227 | (68) | 155 | (60) |  | (43) | 19 | (11) | 22 | (50) |  | (18) | 97(41 | 41.7) |
|  |  |  |  |  |  |  |  |  |  |  |  |  | D 80 | . 922 | 22.81 |

Table 6.4.2.2. Estimates of numbers of 1 SW salmon of Maine origin which would not have been harvested in the Newfoundland-Labrador commercial fisheries, 1984-89, if the fisheries were closed each year on the same day that the fisheries were closed in 1991. The number in parenthesis is the percentage of the total numbers of 1 SW Maine-origin salmon caught in SFA.


Table 7．1 Inventory of Parasites and Diseases of Wild and Reared Salmon in Countries at the West Greeniand and N．E．Atlantic Commissions Areas of NASCO

| Diseases |  | $\mathscr{U}$ |  |  |  | $\begin{aligned} & \text { Q } \\ & \underset{Z}{z} \\ & \underset{i}{z} \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { M } \\ & \text { Z } \\ & \text { Z } \\ & \text { 足 } \end{aligned}$ |  |  |  |  | $\begin{aligned} & z \\ & 3 \\ & 3 \\ & 0 \\ & z \end{aligned}$ | 8 <br>  <br>  <br> 0 <br> 0 | $\begin{aligned} & z \\ & \vdots \\ & \vdots \end{aligned}$ | $\begin{aligned} & \text { z } \\ & \text { 号 } \\ & \text { 岕 } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| A．Viral Diseases <br> VHS <br> IHN <br> IPN <br> Viral papilloma <br> VEN／EIBS <br> Swim bladder tumor <br> B．Diseases of unknown etiology <br> Pancreas Disease <br> ISA <br> Epitheliocystis |  |  |  |  | $\begin{aligned} & 0 \\ & 0 \\ & \mathrm{~F} \\ & \mathrm{~F} \\ & 0 \end{aligned}$ |  | $\begin{aligned} & 0 \\ & 0 \\ & \mathrm{~F} \end{aligned}$ <br> F <br> 0 |  | $\begin{aligned} & 0 \\ & 0 \\ & 0 \end{aligned}$ <br> F | $\begin{aligned} & 0 \\ & 0 \\ & \mathrm{~F} \\ & \mathrm{WF} \\ & \mathrm{~F} \\ & 0 \end{aligned}$ |  | 0 <br> 0 <br> FW <br> FW <br> FW <br> F <br> F | $\begin{aligned} & 0 \\ & 0 \\ & \text { FW } \\ & \text { FW } \\ & \text { F } \\ & \text { F } \\ & \text { F } \end{aligned}$ |  |  |


| Diseases | $\begin{aligned} & \sum_{S}^{3} \\ & 0 \\ & 0 \\ & \text { H } \\ & \end{aligned}$ | $\check{U}$ |  |  |  | 录 |  |  | $\begin{aligned} & \text { 足 } \\ & \underset{y}{4} \\ & \underset{\sim}{0} \end{aligned}$ | $\begin{aligned} & \underset{\sim}{z} \\ & \underset{y y y}{\|c\|} \\ & \underset{y}{c} \end{aligned}$ |  | $\begin{aligned} & z \\ & 3 \\ & 3 \\ & 0 \\ & 0 \\ & Z \end{aligned}$ | 昗 <br>  <br> 0 <br> 0 | $\frac{z}{\sqrt{2}}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| C．Bacterial diseases <br> Aeromonas salmonicida <br> Aeromonas sp．（motile） <br> Renibacterium salmoninarum <br> Yersinia ruckeri <br> Vibrio sp． <br> V．anguillarum <br> Vibrio salmonicida <br> Flexibacter columnaris <br> Flexibacter sp． <br> Pseudomonas sp． <br> Serratia sp． <br> Lactobacillus sp． <br> Mycobacterium sp． |  |  |  |  | FW F F 0 F F F F F |  | $\begin{aligned} & \mathrm{F} \\ & \mathrm{~F} \\ & 0 \\ & 0 \\ & \mathrm{~F} \\ & \mathrm{~F} \\ & 0 \\ & \mathrm{~F} \end{aligned}$ |  | F <br> F <br> WF <br> F <br> F <br> F <br> 0 <br> F <br> F <br> F | WF <br> WF <br> 0 <br> F <br> WF <br> WF <br> 0 <br> F <br> WF <br> 0 |  | FW <br> FW <br> FW <br> F <br> FW <br> FW <br> F <br> FW <br> F | FW <br> FW <br> FW <br> F <br> FW <br> F <br> FW <br> F <br> F |  |  |


| F | $=\quad$ Found in farmed fish． |
| :--- | :--- |
| W | $=$ Found in wild fish． |
| 0 | $=$ Looked for but not found． |
| Blank | $=$ No records． |


| Diseases |  | $\cong$ |  | $$ |  |  | $\begin{aligned} & \overleftrightarrow{U} \\ & Z \\ & \underset{\sim}{4} \\ & \frac{1}{2} \end{aligned}$ |  | $\begin{aligned} & \underset{Z}{\underset{Z}{Z}} \\ & \underset{\sim}{U} \\ & \hline \end{aligned}$ | $\begin{aligned} & \underset{\sim}{2} \\ & \underset{\sim}{y} \\ & \underset{\sim}{3} \end{aligned}$ |  | 2 2 2 0 0 $z$ | $\begin{aligned} & \text { 艺 } \\ & \underset{y}{3} \\ & 0 \\ & 0 \end{aligned}$ | $\frac{z}{\frac{z}{6}}$ | $\begin{aligned} & \text { Z } \\ & \vec{a} \\ & 0 \\ & \overrightarrow{3} \\ & \underset{\sim}{3} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| D. Fungal infections <br> Ichthyophonus hoferi <br> Exophiala salmonis <br> Phoma herbarum <br> Saprolegnia parasitica <br> Saprolegnia sp. <br> Saprolegnia diclina <br> Dermocystidium sp. <br> Paecilomyces farinosus <br> Phialophora sp. |  |  |  |  |  |  | $\begin{array}{\|l} 0 \\ \mathrm{~F} \\ 0 \end{array}$ <br> WF |  | F | $\begin{aligned} & \text { W } \\ & \text { F } \\ & \text { F } \\ & \text { WF } \\ & \text { WF } \\ & 0 \\ & 0 \\ & \text { F } \end{aligned}$ |  | F F <br> FW <br> FW | $\begin{aligned} & \text { W } \\ & \text { F } \\ & \text { FW } \\ & \\ & \text { FW } \\ & \text { F } \\ & \text { F } \\ & \text { F } \\ & \text { F } \end{aligned}$ |  |  |


| Diseases |  | $\mathscr{U}$ | $\frac{\underset{C}{x}}{\substack{x\\}}$ | $\underset{\substack{\mathrm{Z}}}{\underset{y y}{Z}}$ |  | $\begin{aligned} & \text { 忐 } \\ & \underset{B}{\underset{B}{z}} \end{aligned}$ |  |  |  |  |  | $\begin{aligned} & z \\ & z \\ & z \\ & 0 \\ & z \end{aligned}$ | $\begin{aligned} & \text { O} \\ & \mathbb{Z} \\ & E \\ & 0 \\ & 0 \end{aligned}$ | $\frac{z}{\frac{z}{c}}$ | $\begin{aligned} & \text { z } \\ & \stackrel{1}{0} \\ & \text { n } \\ & \vdots \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| E. Protozoan infections <br> Myxobolus neurobius <br> Myxidium truttae <br> Myxidium oviforme <br> PKX organism (probable myxosporid) <br> Ichthyobodo (Costia) necatrix <br> Ichthyophthirius multifiliis <br> Hexamita sp. <br> Trichodina sp. <br> Leptotheca sp. <br> Epistylis sp. <br> Apiosoma sp. <br> Scyphidia sp. <br> Chilodonella cyprini <br> Trichophyra sp. |  |  |  |  |  |  | $\begin{aligned} & F \\ & F \\ & F \end{aligned}$ |  | F <br> F | F <br> F <br> FW <br> F <br> FW <br> F <br> F <br> F |  | W <br> FW <br> FW <br> F <br> F <br> FW <br> FW <br> FW <br> FW <br> FW <br> FW | W <br> W <br> W <br> F <br> FW <br> FW <br> F <br> FW <br> FW <br> F <br> F <br> FW <br> FW <br> F |  |  |


| F | $=$ | Found in farmed fish. |
| :--- | :--- | :--- |
| W | $=$ | Found in wild fish. |
| $\mathbf{0}$ | $=$ | Looked for but not found. |
| Blank | $=$ | No records. |


| Diseases | $\underset{\substack{M \\ M \\ M}}{\substack{M \\ \hline}}$ | $\frac{\varkappa}{U}$ |  |  | $\begin{aligned} & \text { 冗 } \\ & \text { Q } \\ & \underset{\sim}{4} \end{aligned}$ |  |  | $\begin{aligned} & \hat{Z} \\ & \underset{z}{i} \\ & \underset{u}{u} \\ & \frac{0}{0} \end{aligned}$ |  |  | $$ | $\begin{aligned} & z \\ & z \\ & z \\ & 0 \\ & 0 \\ & z \end{aligned}$ | $\begin{aligned} & 0 \\ & z \\ & \vdots \\ & \vdots \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \underset{Z}{Z} \\ & \underset{\sim}{\Sigma} \end{aligned}$ | $\begin{aligned} & \text { z } \\ & \text { M } \\ & 0 \\ & 3 \\ & 3 \\ & 3 \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| F．Monogeneans <br> Gyrodactylus derjavini Gyrodactylus truttae Gyrodactylus salaris Discocotyle sagittata |  |  |  |  |  |  |  |  |  | $\begin{aligned} & 0 \\ & 0 \\ & 0 \end{aligned}$ |  | $\begin{aligned} & \text { FW } \\ & 0 \\ & \text { FW } \\ & \text { W } \end{aligned}$ | $\begin{aligned} & \text { FW } \\ & \text { FW } \\ & 0 \\ & \mathrm{~W} \end{aligned}$ |  |  |
| G．Trematodes <br> Crepidostomum farionis <br> Diplostomum spathaceum <br> Diplostomum sp． <br> Apatemon sp ． <br> Phyllodistomum simile <br> Hemiurus sp． <br> Derogenes sp． <br> Lecithaster sp． <br> Brachyphallus sp． <br> Tetracotyle sp． |  |  |  |  |  |  |  |  |  | W W FW |  | W <br> W <br> W <br> W <br> FW <br> W | W <br> W <br> W <br> W <br> W <br> W <br> W <br> W <br> W <br> W |  |  |


| Diseases | $$ | $\mathscr{U}$ |  | $\begin{aligned} & \text { Z } \\ & \underset{y}{z} \\ & \underset{y y}{z} \\ & \hline \end{aligned}$ |  |  | $\begin{aligned} & \text { U } \\ & \text { Z } \\ & \underset{\sim}{\sim} \\ & \underset{L}{2} \end{aligned}$ |  |  | $\begin{aligned} & \hat{Z} \\ & \underset{y}{z} \\ & \underset{\sim}{u} \end{aligned}$ |  | 2 2 2 0 0 $z$ | 艺 $\vdots$ $\vdots$ $\vdots$ 0 | $\begin{aligned} & Z \\ & \frac{Z}{4} \\ & \stackrel{4}{4} \end{aligned}$ | $\begin{aligned} & \text { 乭 } \\ & \text { 男 } \\ & \text { 会 } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| H．Cestode infections <br> Cyathocephalus truncatus <br> Diphyllobothrium ditremum larvae <br> Diphyllobothrium dendriticum larvae <br> Diphyllobothrium sp．larvae Eubothrium crassum Hepatoxylon sp．larvae |  |  |  |  |  |  |  |  |  | W WF <br> W WF |  | W <br> W <br> W <br> FW | W <br> WF <br> WF <br> W <br> WF <br> W |  |  |
| I．Nematode infections <br> Anisakis sp．larvae <br> Hysterothylaceum sp ．（larvae and adults） <br> Capillaria salvelini <br> Capillaria sp． <br> Metabronema sp． <br> Rhabdochona salvelini <br> Rhabdochona sp． <br> Cystidicola farionis <br> Cystidicoloides sp． |  |  |  |  |  |  |  |  |  | W |  | W <br> FW <br> W | W <br> W <br> W <br> W <br> W <br> W <br> W <br> W <br> W |  |  |


| F | $=$ | Found in farmed fish． |
| :--- | :--- | :--- |
| W | $=$ | Found in wild fish． |
| 0 | $=$ | Looked for but not found． |
| Blank | $=$ | No records． |


| Diseases |  | 든 |  | $\begin{aligned} & \text { 号 } \\ & \stackrel{\rightharpoonup}{2} \\ & \underset{\sim}{2} \end{aligned}$ | $\begin{aligned} & \ddot{\ddot{H}} \\ & \stackrel{0}{4} \\ & \hline \end{aligned}$ | $\begin{aligned} & \sum_{2}^{Z} \\ & \frac{3}{4} \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { w } \\ & \underset{\widetilde{~}}{\substack{w}} \end{aligned}$ |  | $\begin{aligned} & \stackrel{0}{4} \\ & \text { 苞 } \end{aligned}$ |  |  |  |  | $\begin{aligned} & z \\ & \text { z} \\ & \text { in } \end{aligned}$ | $\begin{aligned} & \text { 亶 } \\ & \text { 劵 } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| J．Acanthocephalan infections Neoechinorhynchus rutili Echinorhynchus truttae Pomphorhynchus laevis Acanthocephalus lucii |  |  |  |  |  |  |  |  |  |  |  |  | W W w w |  |  |
| K．$\frac{\text { Crustacean infections }}{\text { Lepeophtheirus salmonis }} \begin{aligned} & \text { Caligus elongatus } \\ & \text { Salmincola salmonea }\end{aligned}$ |  |  |  |  | $\begin{aligned} & \mathrm{FW} \\ & \mathrm{FW} \end{aligned}$ |  | $\begin{aligned} & \mathrm{F} \\ & \mathrm{~F} \end{aligned}$ |  | $\begin{aligned} & \mathrm{F} \\ & \mathrm{~F} \end{aligned}$ | $\begin{aligned} & \mathrm{wF} \\ & \mathrm{wF} \end{aligned}$ |  | $\begin{aligned} & \mathrm{FW} \\ & \mathrm{FW} \\ & \mathrm{FW} \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { WF } \\ & \text { WF } \\ & \mathrm{W} \end{aligned}$ |  |  |
| L．Molluscan infections Margaritifera margaritifera （glochidia） Mytilus edulis |  |  |  |  |  |  |  |  |  | F |  | $\begin{aligned} & \mathrm{FW} \\ & \mathrm{FW} \end{aligned}$ | $\begin{aligned} & \mathrm{F} \\ & \mathrm{~F} \end{aligned}$ |  |  |
| $\begin{array}{\|c\|} \hline \text { M. } \frac{\text { Leech infections }}{\text { Hemiclepsis marginata }} \\ \hline \end{array}$ |  |  |  |  |  |  |  |  |  |  |  |  | w |  |  |


| F | $=$ Found in farmed fish． |
| :--- | :--- |
| W | $=$ Found in wild fish． |
| 0 | $=$ Looked for but not found． |
| Blank | $=$ No records． |

Table 9.1 Production of farmed salmon in the North Atlantic area
(in tonnes round fresh weight), 1980-1991.

| Year | Scotland | Norway | Iceland | Canada | N.Irel. | Faroe Is. | USA | Russia | Ireland | Total |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1980 | 598 | 4153 |  | 11 |  |  |  |  | 21 | 4783 |
| 1981 | 1133 | 8422 |  | 21 |  |  |  |  | 35 | 9611 |
| 1982 | 2152 | 10266 |  | 38 |  | 70 |  |  | 100 | 12626 |
| 1983 | 2536 | 17000 |  | 69 |  | 110 |  |  | 257 | 19972 |
| 1984 | 3912 | 22300 |  | 227 |  | 120 |  |  | 385 | 26944 |
| 1985 | 6921 | 28655 | 91 | 359 |  | 470 |  |  | 700 | 37196 |
| 1986 | 10338 | 45675 | 123 | 672 |  | 1370 |  |  | 1215 | 59393 |
| 1987 | 12721 | 47417 | 490 | 1334 |  | 3530 | 365 |  | 2232 | 68089 |
| 1988 | 17951 | 80371 | 1053 | 3542 |  | 3300 | 455 |  | 4700 | 111372 |
| 1989 | 28553 | 114866 | 1480 | 5865 |  | 8000 | 905 |  | 5068 | 164737 |
| 1990 | 32350 | 157944 | 2800 | 7810 | 60 | 13000 | 2086 | 5 | 8175 | 224230 |
| 1991 | 40593 | 140000 | 2566 | 9395 | 100 | 15000 | 3650 | 0 | 6265 | 217569 |

Table 10.1
Number of microtags, external tags and finclips applied to Atlantic salmon by countries for 1991


Figure 3.2.1.1. Total landings of European stocks in home and distant water fisheries.


UK,IRE,EUR=United Kingdom, Ireland, mainland Europe
RUS,SCAN=Russia, Scandinavia
ICE = Iceland
GRE,FAR,OTH=European component of Greenland, Faroes, others

Figure 3.2.1.2


Scottish MSW Salmon Catches, In Numbers, As Percentage Difference From The 1952-1990 Mean


Scottish 1 SW \& MSW Salmon Catches, In Numbers, As Percentage Difference From The 1952-1990 Mean



Figure 3.3.1.1. Map of Eastern Canada showing Salmon Fishing Areas (SFA).

Figure 3.3.1.2. Percent change between harvest of salmon in 1991 and the average for 1986-1990 (comm) and 1984-89 (rec) in SFAs or zones of Canada. Percent change in 1991 egg depositions from 1986-90 is for specifc rivers.*


Figure 3.3.1.3. Recreational harvest of Atlantic salmon in the wild run rivers of Maine (USA). Limited to 2SW salmon harvest for five rivers: Dennys, E. Machias, Machias, Narraguagus, and Sheepscot.


Figure 3.3.1.4. Commercial landings of Canadian origin salmon in home and distant water fisheries.


NS=Nova Scotia
NB=New Brunswick
QUE=Quebec
NFLD = Newfoundland-Labrador
GRLD $=$ North American component of Greenland

Figure 3.3.1.5. Smoothed $Z$-scores of counts and estimates of small salmon ascending 15 rivers in Canada.


Figure 3.3.1.6. Smoothed Z-scores of counts and estimates of large salmon ascending 14 rivers in Canada.


Figure 3.3.3.1. Indices of smolt survival for Western North Atlantic salmon. Top: Wild smolts Canada, Middle: Hatchery smolts Canada, Bottom: Penobscot River hatchery smolts. Data smoothed by three year means.


Figure 4.1.1.1 The Faroe Exclusive Economic Zone (EEZ).


Figure 4.1.2.1. Catch of salmon in number/10 by statistical rectangle from logbooks in the 1990/1991 season.


Figure 4.1.3.1. Catch per unit of effort ( 1000 hooks) of salmon by statistical rectangle from logbooks in the 1990/1991 season.


Figure 5.1.3.1. Comparison of harvest estimates for Maine-origin (USA) salmon at West Greenland.


Figure 5.1.3.2. Harvest with confidence limits for 1987-90.


Figure 5.1.4.1. Recovery rate of European and North American origin microtags at West Greenland.


Figure 5.1.4.2. North American proportions standardized to 1990.


Figure 5.3.1.1. Schematic diagram of run-reconstruction model for North American Atlantic salmon populations.


Figure 5.3.1.2. Feasible range of Canadian exploitation rates (Eg(i)) and proportion of stock returning from Canada (P) based on estimated ranges of 1SW returns and catch of non-maturing 1 SW salmon in year i.


Figure 5.3.1.3. Feasible range of exploitation rates in Canada (Ec(i)) versus Greenland (Eg(i)) in year $i$ as constrained by estimated ranges of total 2SW returns and harvests of 2SW salmon in Canada in year i+1.
[A and B] Lower and upper bounds for Exploitation rate in Greenland based on upper and lower bounds of grilse exploitation in Canada, respectively. Equations for $A$ and $B$ are presented in Figure 5.3.1.2.


Figure 5.3.1.4. Contour curves of Greenland exploitation rate (Eg(i)) based on maximun values of $\mathrm{Hc}_{\mathrm{nm}, \max }$. Upper and lower bounds of $\mathrm{Eg}(\mathrm{i})$ taken from Figure 5.3.1.3.


Figure 5.3.1.5. Contour curves of Greenland exploitation rate (Eg(i)) based on minimum values of $\mathrm{Hc}_{\mathrm{nm}, \text { min }}$. Upper and lower bounds of $\mathrm{Eg}(\mathrm{i})$ taken from Figure 5.3.1.3.


Figure 5.3.1.6. Estimated pre-fishery abundance (year i) of 1SW salmon of North American origin destined to return as 2SW fish in year (i+1). Estimate includes all salmon regardless of location.


Figure 5.3.1.7. Estimated total abundance of salmon in the West Greenland area prior to the commencement of the fishery. Estimate includes both North American and European stocks and all sea ages.


Figure 5.3.1.8. Estimated range of proportions $(\mathrm{P})$ of North American salmon returning from Canadian fisheries. The proportion returning from Greenland is defined as 1-P-FU where FU is the fraction of the stock unavailable to any fishery.


Figure 5.3.2.1. Predicted number of spawners remaining after fisheries on non-maturing 1SW salmon in Canada (C1) and Greenland (G1) and 2SW salmon in Cananda (C2). Pre-fishery abundance of 1SW salmon destined to return as 2SW spawners is 200,000 fish.


Figure 5.3.2.2. Predicted number of spawners remaining after fisheries on non-maturing 1SW salmon in Canada (C1) and Greenland (G1) and 2SW salmon in Cananda (C2). Pre-fishery abundance of 1SW salmon destined to return as 2SW spawners is 400,000 fish.


Figure 5.3.2.3. Predicted number of spawners remaining after fisheries on non-maturing 1SW salmon in Canada (C1) and Greenland (G1) and 2SW salmon in Cananda (C2). Pre-fishery abundance of 1SW salmon destined to return as 2SW spawners is 600,000 fish.


Figure 5.3.2.4. Relationship between stock and recruitment and stock condition.


Figure 5.3.4.1. Effect of Carlin tag reporting rate and proportion of Maine origin stocks available to the fisheries in Greenland and Canada. Upper line of each panel represents fishery area exploitation with a tag reporting rate adjustment of 2; lower line, reporting rate is unadjusted. Midpoint represents average of the perimeter values.


Figure 6.1.1. Canadian landings of Atlantic salmon, 1974-1991.

Commercial Harvest


Recreational Harvest
Thousands of Salmon


Figure 9.1. Production of farmed salmon in the North Atlantic area.


## APPENDIX 1

## TERMS OF REFERENCE

(C. Res. 1991/2:7:5 The Working Group on North Atlantic Salmon (Chairman: Dr. K. Friedland, USA) will meet in Dublin, Ireland from 5-12 March 1992 to:

1. With respect to Atlantic salmon in each Commission area, where relevant:
a) describe the events of the 1991 fisheries with respect to catches, gear, effort, composition and origin of the catch (including escapees and sea ranched fish), and rates of exploitation;
b) describe the status of the stocks occurring in the Commission area;
c) begin a time series of aggregate estimates of all unreported catches, including those taken in international waters (the latter should be provided separately);
d) specify data deficiencies and research needs;
e) evaluate the by-catches of fish, birds and marine mammals in the salmon drift-net fisheries.
2. With respect to the West Greenland Commission, propose and evaluate methods to estimate:
a) abundance of salmon in the area of the fishery at the time it operates;
b) total abundance of stocks exploited by the fishery wherever they are;
c) possible catch levels based upon maintaining adequate spawning biomass;
d) some index based on the rivers which make a major contribution to the West Greenland fishery.
3. Evaluate the following management measures on the stocks and fisheries occurring in the respective Commission areas:
a) regulations introduced into the Norwegian salmon fisheries in 1989;
b) quota management measures taken in 1990 and 1991 in the Newfoundland and Labrador commercial salmon fisheries.
4. With respect to Atlantic salmon in the Northeast Atlantic Commission and West Greenland Commission areas, provide an inventory of parasites and diseases of wild and reared salmon by country.
5. With respect to Atlantic salmon in the West Greenland Commission area, evaluate the effects which management of the West Greenland fishery has had on stocks in homewaters.
6. With respect to Atlantic salmon in the NASCO area, provide a compilation of microtag, finclip, and external tag releases by ICES Member Countries in 1991.
7. With respect to Atlantic salmon in the West Greenland Commission area, examine historical data on catches and stock composition for the presence of predictable patterns and evaluate the adequacy of sampling programmes to estimate stock compositions by area and time period.

## DOCUMENTS SUBMITTED TO THE WORKING GROUP

1. Dunkley, D.A. Report of the United Kingdom (Scotland) for 1991.
2. Reddin, D.G. Evaluations of trends in North American proportion at Greenland.
3. Reddin, D.G. and L.T. Marshall. Index of abundance for the Greenland Fishery.
4. Reddin, D.G. Identification of North American and European Atlantic Salmon (Salmo Salar L.) caught at West Greenland in 1991.
5. Reddin, D.G. 1991 Database for discrimination at Greenland.
6. Reddin, D.G. and P.B. Short. Lenght, Weight, and Age Characteristics of Atlantic Salmon (Salmo Salar L.) of North American and European Origin Caught at West Greenland in 1991.
7. Reddin, D.G. Environmental conditions in the Northwest Atlantic.
8. M $\phi$ ller Jensen, J. The Salmon Fishery at West Greenland 1991.
9. Russell, I.C., E.C.E. Potter, D.G. Reddin and K.D. Friedland. Recoveries of coded wire microtags from salmon caught at West Greenland in 1991.
10. Potter, E.C.E., L. Kell and D.G. Reddin. The use of a neutral network to distinguish North American and European Salmon (Salmo Salar L.) using scale characteristics.
11. Friedland, K.D., L.W. Stolte, T.F. Meyers and E.T. Baum. Estimated Harvest of USA-Origin 1-SW Salmon in Greenland in 1987-1990.
12. Friedland, K.D., P.J. Rago and E.T. Baum. Carlin Tag returns and harvest estimates of USA Origin Salmon in Greenland, 1967-1991.
13. Stolte, L.W. Development and implementation of a Video exchange program between the Atlantic Salmon producing Countries.
14. Rago, P.J. and K.D. Friedland. Marine exploitation of maine Atlantic Salmon stock.
15. Rago, P.J. Continental run reconstruction models for Atlantic Salmon: application to North America and West Greenland.
16. Rago, P.J., D.G. Reddin and K.D. Friedland. Harvest estimation procedures for USA Salmon stocks: proportional Harvest method.
17. Baum, E.T. Diseases and /or disease agents of Atlantic Salmon in the USA.
18. Anon. Report of the Study Group on the North American Salmon Fisheries.
19. Anon. Report of the Study Group on the Norwegian Sea and Faroes Salmon Fishery.
20. Anon. Report of workshop on salmon assessment methods.
21. Marshall, L.T. Database - Scotia Fundy Region, Canada.
22. Rago, P.J. Methods of Trend Analysis.
23. Anon., 1991. Extract of the report of the Working Group on Pathology and diseases of Marine Organisms.

## APPENDIX 3

## REFERENCES

Anon. 1982. Report of Meeting of the Working Group on North Atlantic Salmon. ICES Headquarters, Copenhagen, 13-16 April 1982. ICES, Doc. C.M. 1982/Assess: 19.

Anon. 1984. Report of Meeting of the Working Group on North Atlantic Salmon. Aberdeen, 20 April - 4 May 1984. ICES, Doc. C.M.1984/Assess:16.

Anon. 1985. Report of Meeting of the Working Group on North Atlantic Salmon. St. Andrews, New Brunswick, Canada, 18-20 September 1984. ICES, Doc. C.M.1985/Assess:5

Anon. 1986. Report of the Working Group on North Atlantic Salmon, Copenhagen, 17-26 March 1986. ICES, Doc. C.M.1986/Assess:17.

Anon. 1987. Report of the Working Group on North Atlantic Salmon. Copenhagen, 21-31 March 1988. ICES, Doc. C.M.1988/Assess: 16

Anon. 1988a. Report of the Working Group on the North Atlantic Salmon. Copenhagen, 21-31 March 1988. ICES Doc. C.M.1988/Assess:16

Anon. 1988b. Report of the Study Group on the North American Salmon Fishery. St. John's, Newfoundland, 1-4 March 1988. ICES, Doc. C.M. 1988/M:4

Anon. 1989a. Report of the Working Group on North Atlantic Salmon, Copenhagen, 15-22 March 1989. ICES, Doc. C.M.1989/Assess:12

Anon. 1989b. Report of the Study Group on the North American Salmon Fishery, Leetown, W.Va. USA, 28 February - 3 March 1989. ICES, Doc. C.M. 1989/M:3

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Anon. 1992c. Report of the Workshop on Salmon Assessment Methodology. Dublin, 2-4 March 1992. ICES, Doc. C.M. 1992/M:8

Anon. 1992d. ICES Compilation of Microtag, Finclip and External Tag Releases in 1991. ICES, Doc. C.M.1992/M:9

Anon. 1992e. Report of the Working Group on Pathology and Diseases of Marine Organisms. Copenhagen, 2-5 March 1992. ICES, Doc. C.M.1992/F:2

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## APPENDIX 4

## RECOMMENDATIONS FROM STUDY GROUPS AND WORKSHOPS

## Recommendations of the Study Group on the Norwegian Sea and Faroes Salmon Fishery

1. The Study Group on the Norwegian Sea and Faroes Salmon Fishery should be re-established as the Study Group on North East Atlantic Salmon Fisheries.
2. The new Study Group should meet in Copenhagen for 4 days immediately prior to the meeting of the Working Group on North Atlantic salmon in 1993 to compile relevant data. [If additional questions are posed the meeting should be extended to 5 days.]
3. Data should be provided to re-run the national run-reconstruction model for 1989 and 1990 smolt releases, and a sensitivity analysis should be prepared.

## Recommendations of the Study Group on North

 American Salmon Fisheries1. The Study Group noted apparent spatial and temporal clustering of the 1991 CWT recoveries from the Canadian fishery and recommended additional analyses to discern broad scale migration patterns.
2. Details of river returns and spawners (by seaage), egg deposition and target egg depositions have now been provided for 19 stocks/stock complexes in the North American Commission Area. The Study Group recommended that data for all stocks be formatted similar to that of Table 2.2.3.1 of this report.
3. Numerous tabulations of catch, effort, run-size and sea-survival have been provided to infer the present status of stocks and recent/past trends. The Study Group encouraged national agencies to consider the adoption of graphic methods to depict measures of central tendency, trends etc., in their data.
4. Little or no progress was reported on several "requirements" itemized in Anon. (1991b). These topics remain relevant to the development and refinement of models to estimate stock abundance and exploitation rates in marine fisheries. The Study Group again recommended:
a. Investigation of stock-recruitment relationships for naturally spawning fish in the Penobscot and Saint John rivers.
b. Examination of forecast models for MSW salmon in an attempt to explain observed recent decreases in the numbers of MSW salmon and increases in the numbers of 1 SW salmon returning to some Canadian and USA rivers.
c. Evaluation of relationships between redd counts and spawning populations and its applicability for North American rivers.
d. Estimation of maturation rates in the Newfoundland-Labrador salmon fishery (for details, see Sec. 5.2, No. 9 in Anon. 1991b).
5. If there is a requirement by ICES for the Study Group to meet in 1993, it is recommended that the meeting occur in Leetown W. VA. prior to the meeting of the Working Group on North Atlantic Salmon.

## Recommendations of the Workshop on Salmon Assessment Methodology

1. It was recommended that contacts established during the workshop should be maintained. When further development or refinements take place, a further workshop could be considered. A further workshop could be considered if a special topic or theme warranted review.

## APPENDIX 5

## NAMES AND ADDRESSES OF PARTICIPANTS




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[^0]:    1 Exploits River
    4 L. Terra Nova River
    9 Gander River
    (a) Bishop's Falls
    (b) Gt. Rattling Brook

    5 U. Terra Nova River
    6 Rocky River
    2 Gander River (Salmon Brook)
    7 Northeast River (Placentia)
    8 Grand Bank Brook
    10 Biscay Bay River
    11 Northeast Brook (Trepassey)
    Middle Brook
    12 Colinet River
    13 Conne River
    Partial counts: not included in means

[^1]:    ${ }_{2}^{1}$ Hatchery and wild origin.
    ${ }_{3}^{2}$ Range of estimates provided for Restigouche and Margaree rivers in 1990 and 1991.
    ${ }_{4}^{3}$ Prediction does not adjust for increased counts resulting from release of MSW
    ${ }_{5}^{4}$ Small salmon spawners/small salmon returns.
    ${ }_{6}^{5}$ Mean value
    ${ }^{6}$ Target for the entire river system (including areas under enhancement).
    ${ }_{8}^{7}$ No targets determined for acid-stressed rivers.
    ${ }^{8}$ Target for complete river.
    9 Incl. a transfer of 124 female spawners to this river.
    10 All size groups combined.

[^2]:    ${ }^{1}$ Data from less than 6 sets.

[^3]:    ${ }^{1}$ Provisional figures.

    + Small catch <0.5 t.
    - No catch.

[^4]:    1 Percent change from 1986-90.
    NA=Not Available.

[^5]:    Percent change from 1984-89, quota started 1990.
    Season closed on all rivers in area.
    No catch statistics collected.

[^6]:    a Allowance of 80 t .
    b Southern Labrador re-opened July 20-July 30.
    c Includes landings in SFA 1 which did not have a quota.

