# ICES AFWG REPORT 2010 

ICES Advisory Committee

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# Report of the Arctic Fisheries Working Group <br> (AFWG) 

## 22-28 April 2010

Lisbon, Portugal /Bergen, Norway

# International Council for the Exploration of the Sea Conseil International pour l'Exploration de la Mer 

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## Executive Summary

## Cod in subareas I and II (Norwegian coastal waters)

The cod in subareas I and II, Norw egian coastal waters was assessed on the basis of a survey time series 1995-2009 as well as catch at age data. This year, a new catch series for recreational and tourist fisheries was presented to the Working Group.

- The stock has varied without a clear trend since 2002. Both the stock biomass and the recruitment are at a low level compared to the first years in the time series.
- Norwegian authorities have proposed a rebuilding plan for this stock, and this plan was tested by the W orking Group.

Cod in Sub-areas I and II (Northeast Arctic) was assessed using XSA with the same settings as in the 2009 assessment.

- The fishing mortality ( $\mathrm{F}_{5}-10$ ) has declined since 2005 and is estimated to 0.28 for 2009. This is the lowest since 1990. Estimated SSB for 2009 is $1,070,000 \mathrm{t}$. This assessment represents $1 \%$ downward revision of the 2009 SSB and a $10 \%$ upward revision of F in 2008.
- The new "hybrid" recruitment model, introduced in 2008, was used, resulting in recruitment at age 3 of 384 million in 2010, 465 million in 2011 and 484 million in 2012.
- The managers introduced a new element in the HCR when setting the TAC for 2010: A lower limit on F $(0.30)$ when SSB is above $B_{p a}$. This amended HCR was tested and found to be consistent with the precautionary approach.
- A catch in 2011 corresponding to the amended HCR is $703,000 \mathrm{t}$. This catch corresponds to a fishing mortality of 0.30 in 2011. SSB is estimated to increase from 1,145,000 $t$ at the beginning of 2010 to $1,488,000 t$ in 2011. Such high SSBs have previously only been observed in the late 1940s. Earlier maturation means that a larger proportion of the total stock is spawners now compared to these early years.

IUU-catches amounted to near 30\% of the international reported catch in 2005 but have since declined and w ere set to zero in 2009.

Haddock in Sub-areas I and II (Northeast Arctic) was assessed using XSA with the same settings as in the 2009 assessment.

- Previously (1950-2000) the fluctuation in the haddock stock have shown strong cyclic pattern caused by spasmodic recruitment, where stock biomass has been dominated by single cohorts. This picture has changed in recent years where three subsequent cohorts (2004-2006) all are very abundant.
- The fishing mortality (F4-7) in the last three years has declined somewhat and is in 2009 estimated to 0.31 . The current assessment estimated the total stock to be about 6 \% lower and SSB $17 \%$ lower in 2009, compared to the previous assessment.
- In the projection RCT3 was used to estimate recruiting year classes from 2007 and onwards. The results indicate that the 2007 and 2008 year classes are below average, while the 2009 year class is above average.
- A catch in 2011 corresponding to the evaluated and agreed HCR is $303,000 \mathrm{t}$. This catch is likely to keep the fishing mortality in 2011 at ap-
proximately 0.31 . SSB is expected to increase considerably until 2012, while the total stock biomass will decrease from 2010 onwards. The 2010 total stock biomass of 1.1 million is the highest observed in the time series, which goes back to 1950 .

The assessment of haddock is uncertain, and XSA is sensitive to settings which can give different perception of long time trend in stock dynamics. However, the short time trends seem to be captured and agree well with results from surveys. Difficulties in estimating initial stock size are additional problems in the forecast.

IUU-catches have been high in recent years, but have since declined and were set to zero in 2009.

## Saithe in Sub-areas I and II (Northeast Arctic)

The last benchmark assessment was done at WKROUND February 2010. The main conclusions of the benchmark assessment were:

- Expand the catch matrix from 3-11+ to 3-15+
- Base the Norwegian trawl CPUE on data from all quarters and from days with $>20 \%$ but $<80 \%$ saithe in the catches
- Split the two tuning series in 2002
- Reduce the shrinkage in the XSA and remove the time tapered downweighting

This resulted in changes in estimated fishing mortality, spawning stock biomass and recruitment, especially in the last part of the time series.

- In the projections the GM age 3 recruitment of 169 million was used for the 2006 and subsequent year classes.
- A catch in 2011 corresponding to the evaluated and implemented HCR is $173,000 \mathrm{t}$. This catch corresponds to a fishing mortality of 0.31 in 2011. SSB is estimated to decrease from $416,000 \mathrm{t}$ at the beginning of 2010 to $357,000 \mathrm{t}$ in 2011.

Difficulties in estimating initial stock size are the major problem in the forecast. This is due to divergent indices of abundance used in the tuning of the XSA, in addition to lack of reliable recruitment estimates. Prediction of catches beyond the TAC year will, to a large extent, be dependent on assumptions of average recruitment.

Beaked redfish (Sebastes mentella) in Sub-areas I and II (Northeast Arctic) was assessed on the basis of available trends in the fisheries and surveys, as there is no accepted analytical assessment for this stock. There are signs of improved recruitment, but the stock is still at a low level and will remain there for a considerable period irrespective of current management actions. No directed fishery is advised.

Golden redfish (Sebastes marinus) in Sub-areas I and II (Northeast Arctic) was assessed on the basis of available trends in the fisheries and surveys. There is no accepted analytical assessment for this stock but the Gadget model was used for the sixth time as an experimental analytical assessment model.

- Since 1993, recruitment of S. marinus has been extremely low,
- commercial data and surveys show consistent declining trends in the spawning biomass,
- the exploratory assessment conducted using the Gadget simulation model covering the period 1986-2009 showed a reduction of the spawning stock
to about $50 \%$ of the level in the early 1990s, and a more severe reduction of the recruitment and the immature stock,
- present available information confirms last year's evaluation of the very poor status of the stock

Greenland halibut in Sub-areas I and II (Northeast Arctic) is in the category "same advice as last year" this year and last year's advice was repeated. Stock trends in recent years indicate a slight increase in stock size. There is no accepted analytical assessment for the time being. It is hoped that the age reading workshop to be held in 2011 will lead to agreement on age reading methodology
According to ToRb, the data on Barents Sea capelin were updated.

## 0 Introduction

### 0.1 Participants

| Asgeir Aglen | Norway |
| :--- | :--- |
| Ricardo Alpoim | Portugal |
| Matthias Bernreuther | Germany |
| Mette Bertelsen | ICES |
| Bjarte Bogstad (Chair) | Norway |
| Vladimir Borisov | Russia |
| Oleg Bulatov | Russia |
| Tatiana Bulgakova | Russia |
| Mikel Casas | Spain |
| Anatoly Chetyrkin | Russia |
| Gjert Endre Dingser | Norway |
| Konstantin Drevetnyak | Russia |
| Anatoly Filin | Russia |
| Age Fotland | Norway |
| Harald Gjeseter | Norway |
| Elvar Halldor Hallfredsson | Norway |
| Daniel Howell | Norway |
| Yuri Kovalev | Russia |
| Sigbjern Mehl | Norway |
| Kjell H. Nedreaas | Norway |
| Dmitry Prozorkevich | Russia |
| Jon Ruiz Gondra | Spain |
| Alexey Russkikh | Russia |
| Oleg Smirnov | Russia |
| Jan Erik Stiansen | Norway |
| Knut Sunnanå | Norway |
| Ross Tallman | Canada |
| Oleg Titov | Russia |
| Natalia Yaragina | Russia |

### 0.2 Locations of the meting

Due to the problems with air traffic in Europe in April 2010 caused by ash from the volcanic eruption in Iceland, only some members of the WG managed to reach the meeting venue in Lisbon. A number of participants, including the Chairman, met in Bergen, while other participants stayed at home at the national laboratories. The meeting was carried out using communication via Internet (e-mail, Sharepoint, the ICES WebEx conference system). We are very grateful to the ICES secretariat for their assistance with use of WebEx. It is not recommended to carry out WG meetings in this way.

### 0.3 Terms of reference

The Arctic Fisheries Working Group [AFWG]: (Chaired by: Bjarte Bogstad, Norway) will meet in Lisb on, Portugal, 22-28 April 2010 to:
a ) address generic ToRs for Fish Stock Assessment W orking Groups (see tablebelow).
b) for Barents Sea capelin oversee the process of providing intersessional assessment.

The assessments will be carried out on the basis of the stock annex in National Laboratories, prior to the meeting. This will be coordinated as indicated in the table below. Material and data relevant for the meeting must be available to the group no later than 14 days prior to the starting date.

AFWG will report by 6 May 2010 (and 7 October 2010 for Barents Sea capelin) for the attention of ACOM.

| FishStock | Stock Name | Advice |
| :--- | :--- | :--- |
| cod-arct | Cod in Subareas I and II (Northeast Arctic) | Update |
| cod-coas | Cod inSubareas I and II (Norweg ian coastal waters) | Update |
| had-arct | Haddock in Subareas I and II (Northeast Arctic) | Update |
| sai-arct | Saithe in Subareas I and II (Northeast Arctic) | Update |
| cap-bars | Capelin in Subareas I and II (Barents Sea),excluding Division IIa west <br> of5 ${ }^{\circ} \mathrm{W}$ | Update |
| ghl-arct | Greenland halibut in Sub-areas I\& II | Same advice <br> as lastyear |
| smn-arct | Redfish Sebastes mentella Subareas I and II | Update |
| smr-arct | Redfish Sebastes marinus Subareas I and II | Same advice <br> as lastyear |

In addition, AFWG has received the following two requests from the Norwegian Ministry of Fisheries and Coastal Affairs:

1 ) According to paragraph 5.1 of the protocol of $38^{\text {th }}$ session of the JNRFC, the parties discussed the possibility of an amendment of the management plan for the Northeast Arctic cod, to secure a more suitable management regime in periods of strong growth or reassessments of the stock size.

The parties agreed to:
a ) a) Establish a management criterion which introduces a minimum fishing mortality rate (F) of 0.30 - effective from 2010.
b ) b) Request the ICES to confirm that the additional criterion is in line with the precautionary approach, and provide future advice according to the revised management plan.

This new management criterion does not apply if the spawning stock biomass falls below $\mathrm{B}_{\mathrm{pa}}$. For further details regarding the management plan and implementation of the new criterion we refer to Annex 14 of the protocol of the $38^{\text {th }}$ session of the Joint Norwegian-Russian Fisheries Commission.

The Norwegian Ministry of Fisheries and Coastal Affairs - representing the Norwegian party in the JNRFC - would like ICES to comment on the agreed upon amendment to the Northeast Arctic cod management plan, as anchored in the protocol and described above.

2: To evaluate whether the adopted rebuilding plan for Norw egian coastal cod is consistent with the Precautionary Approach. If this is not the case, or if the basis for evaluation is unsatisfactory, further advice for modifications or alternative plans is requested. The rebuilding plan is as follows:
"The overarching aim is to rebuild the stock complex to full reproductive capacity, as well as give sufficient protection to local stock components. Until a biologically founded rebuilding target is defined, the stock complex will only be regarded as restored when the survey index of spawning stock in two successive years is observed to be above 60000 tons'• Importantly, this rebuilding target will be redefined on the basis of relevant scientific information. Such information could, for instance, include a reliable stock assessment, as well as an estimate of the spawning stock corresponding to full reproductive capacity

Given that the survey index for SSB does not increase, the regulations will aim to reduce $F^{2}$ by at least 15 per cent annually compared to the $F$ estimated for 2009. If, however, the latest survey index of SSB is higher than the preceding one - or if the estimated F for the latest catch year is less than 0.1- the regulations will be unchanged. Special regulatory measures for local stock components will be viewed in the context of scientific advice. A system with stricter regulations inside fords than outside fiords is currently in operation, and this particular system is likely to be continued in the future. The management regime employed is aiming for improved ecosystem monitoring in order to understand and possibly enhance the survival of coastal cod. Potential predators are - among others - cormorants, seals and saithe.

Generic ToRs for Regional and Species Working Groups
The following ToRs apply to: AFWG, HAWG, NWWG, NIPAG,WGWIDE, WGBAST, WGBFAS, WGNSSK, WGCSE, WGDEEP, WGHMM, and WGANSA.
The working group should focus on:
ToRs a) to h) for stocks that will have advice,
ToRs b) to f) and $h$ ) for stocks with same advice as last year.
ToRs b) to c) and ff for stocks with no advice.
a) Produce a first draft of the advice on the fish stocks and fisheries under considerations and the regional overview according to ACOM guidelines.
b ) Update, quality check and report relevant data for the working group:
i) Load fisheries data on effort and catches (landings, discards, bycatch, including estimates of misreporting when appropriate) in the $\mathbb{N}$ TERCATCH database by fisheries/fleets. Data should be provided to the data coordinators at deadlines specified in the ToRs of the individual groups. Data submitted after the deadlines can be incorporated in the assessments at the discretion of the Expert Group chair;
ii ) Abundance survey results;
iii) Environmental drivers.
iv) Propose specific actions to be taken to improve the quality of the data (including improvements in data collection).
c) Produce an overview of the sampling activities on a national basis based on the INTERCATCH database);
d ) In cooperation with the Secretariat, update the description of major regulatory changes (technical measures, TACs, effort control and management plans) and comment on the potential effects of such changes including the effects of newly agreed management and recovery plans.
e) For each stock update the assessment by applying the agreed assessment method (analytical, forecast or trends indicators) as described in the stock annex. If no stock annex is available this should be prepared prior to the meeting.
f) Produce a brief report of the work carried out by the W orking Group. This report should summarise for the stocks and fisheries where the item is relevant:
i) Input data (including information from the fishing industry and NGO that is pertinent to the assessments and projections);
ii) Where misreporting of catches is significant, provide qualitative and where possible quantitative information and describe the methods used to obtain the information;
iii ) Stock status and 2011 catch options;
iv ) Historical performance of the assessment and brief description of quality issues with the assessment;
v ) Mixed fisheries overview and considerations;
vi ) Species interaction effects and ecosystem drivers;
vii ) Ecosystem effects of fisheries;
viii ) Effects of regulatory changes on the assessment or projections;
g ) Where appropriate, check for the need to reopen the advice in autumn based on the new survey information and the guidelines in AGCREFA
h ) Set MSY reference points (Fmsy and MSY Btrigger) accor ding to the ICES MSY framew ork and following the guidelines developed by WKFRAME.

### 0.4 Unreported landings

In previous years, estimates of unreported landings of cod and haddock have been made separately by Norway and Russia. This year, a report from the NorwegianRussian analysis group dealing with estimation of total catch of cod and haddock in the Barents Sea in 2009 was presented to AFWG (WD13). The report present estimated catches made by Norwegian, Russian and third countries separately. According to that report the total catches of both cod and haddock reported to AFWG are very close (within 1\%) to the estimates made by the analysis group. Thus it was decided to set the IUU catches for 2009 to zero.

It should, however, be noted that there is some disagreement between the Parties in the analysis Group on the interpretation of mandate of the Group and the approach to be used. Mutual inspection of the other Parties' data, has, for instance, not been carried out. Thus one of the Parties has asked the Joint Norwegian-Russian Fisheries Commission for a clarification of how the mandate should be interpreted.

Unreported landings will reduce the effect of management measures and will undermine the intended objectives of the harvest control rule. It is therefore important that management agencies ensure that all catches are counted against the TAC. The

AFWG therefore expects that Norway and Russia will continue the work to secure the necessary quality and accuracy of the catch statistics. Inspections at sea need to be an important part of this work, and Norway and Russia have check-points in their respective economic zones where all fishing vessels have to pass. There are at present, how ever, no such operative check-points for the fisheries in Spitsbergen waters.

### 0.5 Uncertainties in the data

## Catch data

At recent AFWG meetings it has been recognized that there is considerable evidence of both substantial mis-/unreporting of catches and discarding throughout the Barents Sea for most groundfish stocks having taken place (ICES CM 2002/ACFM:18, ICES CM 2001/ACFM:02, ICES CM 2001/ACFM:19, Dingsor WD 132002 WG, Hareide and Garnes WD 142002 WG, Nakken WD 102001 WG, Nakken WD8 2000 WG, Schöne WD4 1999 WG, Sokolov, WD 92003 WG, Ajiad et al. WD18 2005 WG, WD 242004 WG and WD2 2008 WG). In addition to these WDs, Dingser (2001) estimated discards in the commercial trawl fishery for Northeast Arctic cod (Gadus morhua L.) and some effects on assessment, and Sokolov (2004) estimated cod discard in the Russian bottom trawl fishery in the Barents Sea in 1983-2002. This work should be continued, updated and presented annually to the AFWG.
It becomes a problem for the Sebastes mentella assessment that some countries fishing S. mentella in international waters of the Norwegian Sea do not report their catches to NEAFC and ICES. EU-reported catches are, for example, not split by individual countries. Lack of consistency betw een daily reports from the sea to NEAFC and later official reports by delegates to NEAFC is also worrying.

The capelin catch is not considered misreported. Discarding is considered negligible.

## Survey data

While the area coverage of the winter surveys for demersal fish was incomplete in 1997 and 1998, the coverage was normal for these surveys in 1999-2002. In the autumn 2002, 2006 and winter 2003, 2007 however, surveys have again been incomplete due to lack of access to both the Norwegian and Russian Economic Zones. This affects the reliability of some of the most important survey time series for cod and haddock and consequently also the quality of the assessments. In some years, the permission to work in the Norwegian and Russian Economic Zones, respectively, has been received so late that the work has been severely hampered, e.g., the Russian survey in autumn 2003 and 2006. There is no acceptable way around this problem except asking the Norwegian and Russian authorities to give each other's research vessels full access to the respective economical zones when assessing the joint resources, as, e.g., was the case for Norwegian w inter surveys in 2004-2005 and 2008-2010.

From 2004 onwards, a new joint Norwegian-Russian survey has been conducted in August-September. This is a multi-purpose survey termed an "ecosystem survey" because most part of the ecosystem is covered; including a bottom trawl survey and an acoustic survey for the all species, witch available for assessment, include not commercial species. Ongoing work is considering the performance of these new index series for inclusion in the assessment of cod and haddock, and they seem to be fairly consistent with the other series available (WD20). The survey is also utilised in the assessment of redfish and Greenland halibut. However, this survey may be discontinued or downscaled for economical reasons. This is highly regrettable, since
this survey has been shown to be valuable for sampling of synoptic ecosystem information, cover the all area of fish distribution in the Barents Sea, and addition data on demersal fish, which could prove valuable in future inclusion of more ecosystem information in the fish stock assessments.

## Age reading

In 1992, PINRO, Murmansk and IMR, Bergen began a routine exchange program of cod otoliths in order to validate age readings and ensure consistency in age interpretations (Yaragina et al. 2009b, AFWG 2008, WD 20). Later, a similar exchange program has been established for haddock, Greenland halibut and capelin otoliths. Once a year (for capelin every second year) the age readers have come together and evaluated discrepancies, which are seldom more than 1 year, and the results show an improvement over the time period, despite still observing discrepancies for cod in the magnitude of $15-30 \%$. An observation that is supported by the results of a NEA cod otolith exchange between Norway, Russia and Germany (Høie et al. 2009, AFWG 2009, WD 6). 100 cod otoliths were read by 3 Norwegian, 2 Russian and 1 German reader, reaching nearly $83 \%$ agreement (coefficient of variation $8 \%$ ). The age reading comparisons of these 100 cod otoliths show that there are no reading biases between readers within each country. However, there is a clear trend of bias between the readers from different countries, Russian age readers assign higher ages than the Norwegian and German age readers. This systematic difference is a source of concern and is also discussed in Yaragina et al. (2009b). This seems to be a persistent trend and will be revealed in the following annual otolith and age reader exchanges.

A positive development is seen for haddock age readings showing that the frequency of a different reading (usually $\pm 1$ year) has decreased from above 25\% in 1996-1997 to about $10 \%$ at present. The discrepancies are always discussed and a final agreement on the exchanged cod and haddock otoliths is at present achieved for all otoliths except ca. $2-5 \%$. To determine the effects of changes in age reading protocols between contemporary and historical practices, randomly chosen cod otolith material from each decade for the period 1940s-1980s has been re-read by experts (Zuykova et al. 2009). Although some year-specific differences in age determination were seen between historical and contemporary readers, there was no significant effect on length at age for the historical time period. A small systematic bias in the number spawning zones detection was observed, demonstrating that the age at first maturation in the historic material as determined by the contemporary readers is younger than that determined by historical readers. The difference was largest in the first sampled years constituting approximately 0.6 years in 1947 and 1957. Then it decreased with time and was found to be within the range of 0.0-0.28 years in the 1970-1980s. The study also shows that cod otoliths could be used for age and growth studies even after long storage.

The exchange meeting in 2009 (WD14), found that the percent disagreement between the PINRO and IMR readings have stabilized in recent years at around $20 \%$ for cod, and around $10 \%$ for haddock, which suggests that annual meetings are not necessary. For the future meetings will be bi-annual, while otolith exchange will take place annually.

The otoliths of Greenland halibut are not easy to read especially for older fish. Consequently the readers have difficulties in inter preting real age zones when the fish become older than 5 years (e.g., AFWG 2005, WD 8). Previous comparative readings among three Norwegian age readers, and also between Russian and Norwegian age readers show good agreement and low CV. However, even with acceptable between
reader precisions, there are strong evidences of low accuracy of the age estimates displayed by IMR (Norway). Since 2006, validation work has been continued (Albert et al. 2009) and the Norw egian age readings have been done using the new approach described in the AFWG 2006 report. The validation work continues and in the future the historic time series might eventually be converted to the new age understanding. However, this work is very time consuming and it is difficult to estimate when a full assessment can be conducted using the new approach.

This has caused that only the recent Russian age readings provided by PINRO have been comparable with the historic data series and used for "illustrative" assessment in 2006-2010. It should be noted that VNIRO (Russia) consider that traditional age readings are valid for fish up to 60 cm length (Kuznetsova, WD 25).

An ICES W orkshop on Greenland halibut age reading will take place in February 2011. Hopefully, during this workshop scientists from different institutes will get an agreement on Greenland halibut growth rate.

For capelin otoliths there is a very good correspondence between the Norw egian and Russian age readings, with a discrepancy in less than $5 \%$ of the otoliths. An international (Russia, Norway, Iceland, Canada) age reading w orkshop on capelin was conducted in May 2009 (WD 1). Otoliths from 20 samples ( 390 otoliths) where discussed. Some of these samples had been exchanged earlier, according to the program of annual otolith exchange between Norway and Russia. Other samples were read for the first time during the workshop, including samples from Iceland and Newfoundland.

For some of the samples, a very high agreement was reached after the initial reading by the different experts. In other cases, some disagreement was evident after the first reading. After the initial reading, the results were analysed. The otoliths that caused disagreement were read again and discussed among the readers. After discussion about the reasons for disagreement, some readers wanted to change their view on some of the otoliths. When the samples were read once more, the agreement was 95 \%.

It was concluded that experts from all laboratories normally interpret capelin otoliths equally. Difficult otoliths are sometimes interpreted differently, but these samples are few, and should not cause large problems for common work on capelin biology and stock assessment. All participants noted the great value of conducting joint work on otolith reading, and it was decided to continue the programme of capelin otolith exchange and to involve the labs at Iceland and New foundland in the exchange program. Readers from Norway and Russia will continue to meet at Workshops every second year. Readers from all labs involved will meet less frequently. Details will be discussed and decided by correspondence.

From 2009 onwards, an exchange of Sebastes mentella otoliths is conducted annually between the Norw egian and Russian laboratories.

## Sa mpling error - catch and survey data

Estimates of sampling error are to a large degree lacking or are incomplete for the input data used in the assessment. How ever, the uncertainty has been estimated for some parts of the input data:

## Catch data

For the Norwegian estimates of catch at age for cod and other demersal species methods for estimating the precision have been developed, and the work is still in
progress (Aanes and Pennington 2003, Hirst et al. 2004, Hirst et al. 2005). The methods are general and can in principle be used for the total catch, including all countries' catches, and provide estimates both at age and at length groups. Typical error coefficients of variation are in the range $5-40 \%$ depending on age and year. It is evident that the estimates of the oldest fish are the most imprecise due to the low numbers in the catches and resulting small number of samples on these age groups. From 2006 onwards, the Norwegian catch at age in the assessment has been calculated using the method described by Hirst et al. (2005).
Aging error is another source of uncertainty, which causes increased uncertainty in addition to bias in the estimates: An estimated age distribution appears smoother than it would have been in absence of aging error. Some data have been analysed to estimate the precision in aging (Aanes 2002). If the aging error is known, this can currently be taken into account for the estimation of catch at age described above.

For capelin, the uncertainty in the catch data is not evaluated. The catch data are used, however, only when parameters in the predation model are updated at infrequent intervals, and the uncertainty in the catch data is considered small in comparison with other ty pes of uncertainties in the estimation.

## Survey data

For the Barents Sea winter survey, the sampling error is estimated per length group, but not per age group. Since the ages are sampled stratified per length groups in this survey, it is not straightforward to estimate the sampling error per age group. However, this is possible by for example using similar methods as for the catch data (see Hirst et al. 2004).

The capelin stock is estimated at the August-September survey. After the survey became a multipur pose survey in 2004, there is a possibility that the amount of trawl catches directed on capelin acoustic registrations has been less than before, as the total number of trawl stations increased. The effect of this on the quality of the capelin estimate has not been quantified. The survey coverage is considered adequate. The uncertainty in the survey has been evaluated by resampling (Tjelmeland 2002), and used as basis for the CV (0.2) chosen for the survey uncertainty in the tool used for calculating the effect of the catch (CapTool) on the spawning stock.

Work on quantifying uncertainties also for other input data sets should be encouraged.

## Sampling effort - commercial fishery

Concerns about commercial sampling: The main Norwegian sampling program for demersal fish in ICES areas I and II has been port sampling, carried out on board a vessel travelling from port to port for approximately 6 weeks each quarter. A detailed description of this sampling program is given in Hirst et al. (2004). However, this program was, for economic reasons, terminated 1 July 2009. Although sampling by the 'reference fleet' and the Coast Guard has increased somewhat in recent years, this change seems to have increased the uncertainty in the catch-at-age estimates (WD6). For the 2009 data, the effect is strongest for saithe, where the fishery is fairly evenly distributed by quarters. Cod and haddock are mainly fished in the first half of the year, so the effect of the change will for those stocks show up much stronger in the 2010 data. Nevertheless, there are already concerns that the commercial sampling could become so poor that analytical assessments cannot be made in the future. The split between coastal cod and NEA cod will affected by this, but no analysis of this is yet available.

The methodological ICES workshops WKACCU (ICES CM 2008/ACOM:32), WKPRECISE (ICES CM 2009/ACOM:40) and WKMERGE (ICES CM 2010/ACOM:40) w ere all dealing with different aspects of catch sampling and the need for a more proper, robust and transparent sampling design for countries involved in catch sampling. The workshops have provided valuable general knowledge in how such catch sampling programs can be designed and the reports are beneficial for countries aiming to improve the current situation.

As most stock -assessment models used at present in ICES (such as standard VPA and the XSA) work with the assumption that the Catch-At-Age data are unbiased, and know exactly, it seems very important to actually be able to assess if this assumption is reasonable by measuring the accuracy of the estimated catch-at-age based on data from sampling programs. Some of the recommendations from different assessment working groups are further related to assessment of the quality of different estimates such as catch-at-age data. To be able to give validation on the data quality it is crucial that the sampling program is set up in a transparent, statistical sound way. Stock assessments need proper sampling designs and estimation processes that are well documented.

### 0.6 Climate included in advice of NEA cod

For the third time climate information has been applied in the advice from AFWG. In this year's assessment ecosystem information was directly used in the projection of NEA cod. A combination of regression models, which is based on both climate and stock parameters, were used for prediction of recruitment at age 3 .
In addition, temperature is part of the NEA cod consumption calculations that goes into the historical back-calculations of the amount of cod, haddock and capelin eaten by cod.

### 0.7 Proposals for status of assessments in 2011-2012

The AFWG propose to set the following status for assessments for each stock:

| FishStock | Stock Name | Advice <br> in <br> 2011* | Previous <br> benchmark | Next <br> benchmark |
| :--- | :--- | :--- | :---: | :---: |
| cod-arct | Cod in Suba reas Iand II (Northeast Arctic) | Update | - | - |
| cod-coas | Cod in Subareas Iand II (Norwe gian coastal waters) | Update | - | - |
| had-arct | Haddock in Subareas Iand II (Northeast Arctic) | Update | - | 2011 |
| sai-arct | Saithe in Subareas Iand II (Northeast Arctic) | Update | WKROUND <br> 2010 | - |
| cap-bars | Capelin in Suba reas Iand II (Barents Sea), <br> excluding Division IIa west of 5\%W | Update | WKSHORT <br> 2009 | - |
| ghl-arct | Greenland ha libut in Sub-a reas I \& II | Same <br> advice <br> as last <br> year | - | - |
| smn-arct | Redfish Sebastesmentella Subareas Iand II | Same <br> advice <br> as last <br> year | - | 2012 |
| smr-arct | Redfish Sebastesmarinus Subareas Iand II | Same <br> advice <br> as last <br> year | - | 2012 |

A benchmark assessment will be planned for Greenland halibut after the age reading workshop, which will be held in February 2011. Such a benchmark assessment should also include the other Greenland halibut stocks.

### 0.8 ICES Quality Handbook

Quality Handbooks for all stocks except Barents Sea capelin are presented in this report as annexes (no. 2-8). For capelin, the stock annex is being updated following the comments made during WKSHORT in 2009 and will be ready before the capelin assessment in autumn 2010. The stock annex for saithe has been updated after the benchmark at WKROUND 2010. For S. mentella, some information on the fishery in International waters in the Norwegian Sea has been added.

### 0.9 InterCatch

The assessment of NEA cod, haddock and saithe was based on output from InterCatch. In the future, AFWG will consider using Intercatch also for the other stocks. It was noted that Intercatch at present does not allow for catches of morethan one stock of a given species in a given area (e.g. Coastal cod and Northeast Arctic cod in ICES area IIa).

### 0.10 MSY-related reference points and advice

## Summary

The AFWG has no difficulty in principle with moving to an MSY based fishery, and considers this to be a valuable extension to the existing precautionary-based approach. How ever we note that the ICES advice for conducting such assessments has only been available recently (and indeed may still be subject to change). We feel that conducting MSY assessments is an involved and complex task, which requires a consideration of the management rule as a whole, and not merely "target $\mathrm{F}^{\prime}$. As such we feel that insufficient time has been available to conduct such assessments at the 2010 WG. The volcano-related travel difficulties that affected this WG have further reduced the time available. We present below the background to our conclusions, and highlight the work that has already been done which could lead to MSY advice in future years, together with the areas that have been identified as requiring detailed consideration for each stock. We would also note that the stocks covered by the AFWG are managed by the Russian and Norwegian governments, neither of whom has requested a move to MSY-based advice in 2011. We believe, in keeping with the view of the Norw egian government, that a move to MSY advice is valuable, but that such a change needs to be well thought out and planned, and not rushed through without due consideration. This is especially important in the AFWG context given that successful management plans are in place for the most commercially important species, and we would be reluctant to provide hasty and under-researched advice that could jeopardize the current successful management of these stocks.

## Background

The generic ToR h) says: Set MSY reference points (Fmsy and MSY Btrigger) according to the ICES MSY framew ork and following the guidelines developed by WKFRAME. In general terms, ICES is aiming at changing the basis for its advice from $\mathrm{F}_{\mathrm{pa}}$ - $\mathrm{B}_{\mathrm{pa}}$ to $\mathrm{Fmsy}^{\prime}$, combined with a trigger spawning biomass ( $\mathrm{B}_{\text {trigger }}$ ). The significance of $\mathrm{B}_{\text {trigger }}$ is that, if a stock is assessed to be below this level, the F for the advice is reduced linearly with SSB.

WKFRAME has given guidelines for calculating FMSY and MSY Btrigger. Also, AFWG has been requested by the ICES secretariat to provide catch options according to FMSY/ MSY Btrigger ${ }^{\text {as well as catch options in accor dance with the adopted EU plan to move }}$ stepwise towards these reference points in such a way that they are reached in 2015. However the complete set of guidelines from WKFRAME was not available until just before the start of the meeting, giving little time for consideration of MSY issues.

Also, in early May 2010 there will be an advisory group meeting on MSY advice (ADGMSY) which will further consider how to incorporate MSY-based approaches into the ICES advice giving process.

AFWG specific issues
In contrast to some other areas, many of the major stocks in AFWG are currently at or near historical maximums, and are successfully managed by existing harvest control rules. There is therefore a desire among both scientists and managers to be cautious in moving away from what has proved to be highly successful management regimes

The stocks assessed by AFWG are managed either by Norway alone (coastal cod and saithe) or through the joint Norwegian-Russian Fisheries Commission (NEA cod and haddock, S. marinus and S. mentella, Greenland halibut, Barents Sea capelin).

In a letter sent to ICES in April 2010, The Royal Norwegian Ministry of Fisheries and Coastal Affairs states that
"...When management authorities request advice according to a management plan which ICES considers being consistent with the precautionary approach, ICES provides such advice. This is the existing situation for the majority of stocks managed by Norway in cooperation with other parties, and Norway has not signaled that any of these plans have yet ceased to exist.

Norway will, in collaboration with the relevant partners, evaluate which revisions are necessary to ensure that the long term management plans provide for maximum sustainable yield. To this end we would welcome any information ICES may have to guide us in the right direction. But as the existing management plans still remain in force, such new information should be given as information or catch options.

Furthermore, as there is a need to anchor the MSY-concept stronger and discuss the short-term consequences amongst the relevant management authorities, I believe it is premature to change the default advisory framework in the advices for 2011."

AFWG has not received any requests from the Russian Federation on the transition to MSY-based advice. AFWG has been informed by ICES that for stocks for which there are agreed management plans the advice for 2011 should be given in accordance with those management plans.

In addition it should be noted that the way the AFWG was carried out this year, with people distributed around Europe, limited the amount of work that could be carried out by the WG.

## MSY-related studies for AFWG stocks

Although we have not been able to give MSY advice during the time period of this meeting, it should be noted that for some stocks, a notable amount of MSY-related studies have already been carried out. This work provides the basis on which the WG could move towards giving MSY based advice, if required by the Norwegian and Russian governments. The AFWG stocks can for the purpose of calculating MSY reference points be divided into 4 groups:

- Stocks for which there is an accepted analytical assessment and an agreed HCR: NEA cod, haddock, saithe
- Stocks for which there are catch-at-age data and reasonable confidence in age readings, but no accepted assessment: Coastal cod, S. marinus, S. mentella
- Stocks for which the age reading methodology is under revision (Age Reading W orkshop to be held in 2011): Greenland halibut
- A short-lived stock with survey-based assessment and an agreed HCR, and for which a single-species MSY is meaningless since predation from cod and other predators is much larger than the fishery: Barents Sea capelin.

For NEA cod, haddock and saithe, there is an accepted analytical assessment and an agreed HCR. All HCRs are similar: $\mathrm{F}=$ constant above $\mathrm{B}_{\mathrm{pa}}$, and F is reduced linearly from this value at $\mathrm{B}_{\mathrm{p}}$ to 0 at $\mathrm{SSB}=0$. In addition there is a constraint on annual variation in TAC (cod: $10 \%$, saithe, $15 \%$, haddock: $25 \%$ ). This constraint is suspended when SSB is below $\mathrm{B}_{\mathrm{pa}}$. For cod and saithe, the anticipated stock development 3 years into the future is taken into account when calculating the TAC. For all stocks, longterm simulations (100 years) using a detailed biological model with stochastic stock/recruitment and density-dependent growth and maturation were used to evaluate whether the HCR is precautionary. For cod, cannibalism was also included in the model. Such simulation models seem to be appropriate to use for MSY studies of these stocks, rather than calculating MSY based on $Y / R$ and SSB/R analyses.
For cod, Kovalev and Bogstad (2005) found that FMSY is in the range 0.25-0.60, where the yield curve is fairly flat (yield in this range within about $80 \%$ of maximum yield), the exact shape is dependent on the biological model used (density-dependent or not, choice of cannibalism model etc.). It should also be noted that Kovalev and Bogstad (2005) found that shifting the exploitation pattern one age group upward would increase the yield, this finding is consistent with other studies (see Kvamme and Bogstad 2007 and references therein). Skagen (2010 WKFRAME WD) found similar results. For this stock, and several other cod stocks, WGSAM (ICES 2008) found that the high yields predicted at low F by single-species models are almost certainly unrealistic, as these will be 'eroded' by predation pressure and density-dependent growth reductions. For NEA cod, using the $\mathrm{SSB} / \mathrm{R}$ at $\mathrm{F}_{01}$ and mean recruitment when the SSB is above $\mathrm{Blim}_{\text {lim }}$ (Figures from 2007 advice report used by ICES WGSAM 2008), gives a SSB of 4.9 million t . This is about four times the historical maximum of 1.2 million t , so $\mathrm{F}_{0.1}$ should not be considered a candidate $\mathrm{Fmsyr}_{\text {reference point. Also for } \mathrm{F}_{\text {max }} \text { values }}$ considerably above the historical maximum were obtained (Section 3). For this cod stock much work has also been done on estimating fecundity and thus total egg production (see recent overview in Morgan et al. 2009) which may affect both fishing mortality reference points and biomass reference/trigger points. This body of work neatly encapsulates the idea that blindly running different values of F through simulations models without considering the wider issues can give results that are highly misleading, and could risk damaging the currently successful management of this stock.

For haddock and saithe, MSY information can similarly be derived from simulations done during the evaluation of whether the HCR for these stocks are precautionary (see AFWG 2006 for haddock and AFWG 2007 for saithe). The yield vs. F curve is rather flat on the top for both stocks. Also, for both stocks, the biological model should bere-visited before any MSY reference points for advisory use are calculated. The reason for this for haddock is the recent strong recruitment and following all-
time high biomass level, which may alter our perception of the stock dynamics. For saithe, the PA reference points were recalculated at AFWG 2010 due to the extended age range to be used in the assessment (seeWKROUND 2010), and this may also alter our perception of the stock dynamics and require new simulations to be made. For saithe we advise not to change the numerical $\mathrm{F}_{\mathrm{pa}}$ and $\mathrm{B}_{\mathrm{pa}}$ values used in the HCR this year, but rather revisit the HCR next year and evaluate it both from a precautionary and MSY point of view.

For the redfish stocks, there are no biomass-based reference or trigger points or management plans. Development of U-type (survey based) biomass reference points has been supported by ACOM. F-based reference points for $S$. mentella were estimated by WKFRAME. For both stocks, AFWG chose not to suggest any of these as Fmsy. One reason for this is that the shape of the growth curve at older ages is uncertain, and this shape strongly affects the yield calculated at low fishing mortalities. It should also be noted that current exploitation rates are well above $\mathrm{F}_{\mathrm{pa}}$, and thus well above the lower Fmsy levels. It is of questionable utility to provide $\mathrm{Fmsy}_{\mathrm{m}}$ estimates for fisheries in which even the more modest $\mathrm{F}_{\mathrm{pa}}$ targets cannot be met.

For coastal cod, there are reasonable data for weight and maturity at age so F-based reference points have been calculated, but no reliable stock/recruitment data. For this stock we have evaluated the proposed rebuilding plan this year. This plan is not linked to a TAC-based or effort-based management, but MSY studies will give knowledge on whether the rebuilding plan is appropriate. Again, providing Fmsyvalues is not urgent given that a rebuilding process is required before the stock can even approach MSY levels.

For Greenland halibut, the managers have agreed upon a fixed yearly quota for the period 2010-2012, and the advice should not be updated this year. The calculation of MSY reference points for this stock will be postponed until after the upcoming age reading workshop in 2011.

For capelin, the agreed HCR is that with $95 \%$ probability, at least 200000 tonnes (Blim) should be allowed to spawn (see Chapter 9 for details). There is no $B_{p a}$ and no Fbased reference points. MSY has been investigated by Tjelmeland (2005), using the multispecies model Bifrost. He found that the MSY reference point of capelin (target SSB) depends markedly on the harvesting strategy chosen for cod and herring, which both have strong biological interactions with capelin. Thus, calculating a singlespecies MSY for capelin is meaningless. The capelin MSY could be calculated given on the agreed HCRs for cod and herring, and one could then investigate whether the MSY for capelin would change considerably if the harvesting strategies for cod and herring vary e.g. within the intervals corresponding to yields $>80 \%$ of the MSY for herring and cod. In the MSY concept paper (WD to WKFRAME 2010), it was stated that the framework outlined there for calculating MSY-based reference points was not applicable for short-lived stocks with a target escapement strategy. WKFRAME did not touch upon this issue.

## General comments on the MSY approach

AFWG also has some comments on the MSY approach in addition to the contents of the WKFRAME report. Most of these are taken from the final report of the EU UNCOVER project (2006-2010).

It should be noted that MSY should by definition mean the maximum sustainable yield that can be obtained from a given stock. An approach that merely involves varying the target F and BMSY trigger within existing management rules will not, in
general, give a MSY fishery. We would also be concerned at an approach that focused only on SSB; when looking at e.g. carrying capacity and not only reproductive potential, total stock biomass (TSB) is just as relevant as SSB and should also be considered in the analysis.
The first point that needs to be made explicit in each and every MSY management rule is that MSY management does not replace the precautionary approach to fisheries management, rather it incorporates and extends it. A management rule that leads to long term reductions of the stock will also lead to reductions in the catches, and is thus by definition not a MSY strategy. A logical consequence of this is that a MSYbased rule should include definitions of where the stock is considered to be at risk of being depleted (i.e. of causing recruitment overfishing), and what remedial action should be taken in this case. This is especially important given the increased uncertainties involved in MSY assessments. MSY fishing should be considered to be "precautionary plus", incorporating and extending the precautionary approach to fisheries, and retaining precautionary biomass limits.

The recommendation, in the absence of an estimated Fmsy or lack of a stockrecruitment relationship is to utilise $\mathrm{F}_{\text {max }}$ or $\mathrm{F}_{0.1}$ as a proxy does not appear to be justified. $\mathrm{F}_{\text {max }}$ is at present determined by ICES WGs ignoring density dependent effects on growth and mortality (including cannibalism) making its utilization questionable. Additionally it is usually very hard to determine the exact value of $\mathrm{F}_{\max }$ in any given model simulation, as curves tend to be very flat topped. Taking a point where the upward slope tends to the asymptote (such as F.1, the point at the slope is $10 \%$ of the maximum), could seem like a good alternative to Fysy in terms of yield, while being precautionary in terms of the stock dynamics. However, for several stocks, combining $\mathrm{Y} / \mathrm{R}$ at $\mathrm{F}_{0} 1$ with average recruitment for spawning stock size above Blim would give a stock size way above what has been observed (see e.g. ICES C.M. 2008/RMC:06) and thus the yield and biomass indicated by such calculations may not be realistic to reach.

We suggest that the default approach to calculating Fmsy should be to base it on simulations of long-term stock dynamics incorporating stock recruitment relationships, density dependent growth and mortality, including uncertainty, environmental issues and possible multispecies effects. Work is therefore required which effects are likely to be of significance for a given stock (environment change, multi-species, mixed fishery, ...), and how to include these in the simulations. These simulation results should then also be used to deduce the time intervals for re-assessing target F and B's. It may be questionable whether it is justified to calculate point estimates of FMSY, giving a range for which the yield is within $80-90 \%$ of the maximum yield (taking into account model uncertainty, choice of length of time series in calculation etc.) could be more appropriate.

### 0.11 Recommendations

AFWG has two recommendations:
A benchmark meeting for all redfish stocks should be held in 2012.
A workshop on methods estimating recruitment for Northeast arctic cod should be held before the AFWG meeting in 2011.

### 0.12 Time and place of Next Meeting

The Working Group proposes to meet next time in Hamburg in the period 5-11 May (alternatively 28 April-4 May) 2011.

The aim of this chapter is to identify important ecosystem information influencing the fish stocks, and further show how this knowledge may be implemented into the fish stock assessment and predictions. There has been a steadily development in this aspect over the last few years and the work is still in a developing phase. Hopefully, the gathering of information on the ecosystem in this chapter will lead to a better understanding of the complex dynamics and interactions that takes place in the ecosystem, and also participate in the development of an ecosystem based management of the Barents Sea.

The ecosystem approach to management is variously defined, but in principle it puts emphasis on a management regime that maintains the health of the ecosystem alongside appropriate use of the marine environment, for the benefit of current and future generations (Jennings, 2004).

Along with fishery, changes in the Barents Sea ecosystem are mainly caused by variations in the ocean climate. A warm period is characterized of increased impact of warm Atlantic water in the Barents Sea contributes to advection of zooplankton, faster growth rate in fish and emergence of abundant year classes (Dalpadado et al. 2002). A cold period is, conversely, characterized by reduced primary biological production in the Barents Sea and emergence of weak year classes of commercial species. Climatic conditions govern the formation of primary biological production and feeding conditions for fish, as well as the survival of their offspring. In addition, inter-species trophic relations are an important factor that influences the abundance dynamics of commercial species.
Movement towards an ecosystem approach to the fishery management in the Barents Sea should include (Filin and Røttingen, 2005):

- More extensive use of ecosystem information in the population parameters applied in assessment and prognosis,
- Expansion of the use of multi-species models for fishing management.

This chapter is in general based on a preliminary version of the 2009 update (Stiansen et al.,WD23) of the "Joint Norwegian-Russian environmental statutes 2008, report on the Barents Sea Ecosystem" (Stiansen et al., 2009), affiliating more than 100 scientists from 24 institutions in Norway and Russia. This report is the successor to the "Joint PINRO/IMR report on the state of the Barents Sea ecosystem in 2007, with expected situation and considerations for management" (Stiansen and Filin, 2008). Text, figures and tables taken from these reports (i.e. Stiansen et al., 2009, or Stiansen et al, WD23) are in general not further cited in this chapter.

### 1.1 General description of the Barents Sea ecosystem (Figure 1.1, Tables 1.1-1.7)

## Geographical description

The Barents Sea is on the continental shelf surrounding the Arctic Ocean. It connects with the Norwegian Sea to the west and the Arctic Ocean to the north. Its contours are delineated by the continental slope between Norway and Spitsbergen to the west, the top of the continental slope towards the Arctic Ocean to the north, Novaya Zemlya archipelago to the east, and the coasts of both Norway and Russia to the south
(Figure 1.1). It covers an area of approximately 1.4 million $\mathrm{km}^{2}$, has an average depth of 230 m , and a maximum depth of about 500 m at the western end of Bear Island Trough (Figure 1.1). Its topography is characterized by troughs and basins ( $300 \mathrm{~m}-$ 500 m deep), separated by shallow bank areas, with depths ranging from $100-200 \mathrm{~m}$. The three largest banks are Central Bank, Great Bank and Spitsbergen Bank. Several troughs over 300 m deep run from central Barents Sea to the northern (e.g. Franz Victoria Trough) and western (e.g. Bear Island Trough) continental shelf break. These troughs allow the influx of Atlantic waters to the central Barents Sea.

## Climate

The general pattern of circulation (Figure 1.1) is strongly influenced by this topography, and is characterised by inflow of relatively warm Atlantic water, and coastal water from the west. This current divides into two branches: 1) a southern branch that flows parallel to the coast and eastwards towards Novaya Zemlya; and 2) a northern branch that flows into the Hopen Trench. The Coastal Water has more freshwater runoff and a lower salinity than the Atlantic water; it also has a stronger seasonal temperature signal. In the northern region of the Barents Sea, fresh and cold Arctic waters flow from northeast to southwest. Atlantic and Arctic water masses are separated by the Polar Front, which is characterised by strong gradients in both temperature and salinity. There is large inter-annual variability in ocean climate related to variable strength of the Atlantic water inflow, and exchange of cold Arctic water. Thus, seasonal variations in hydrographic conditions can be quite large.

## Bacteria and phytoplankton

In the biogeochemical cycles of the ocean, a multitude of processes are catalyzed by Bacteria and Archaea, and the functioning of these cycles in the Barents sea do not differ qualitatively from those at lower latitudes. Both bacteria and viruses show highly variable abundance in the Barents Sea, and in general, the dynamics of these groups in this area do not differ from other parts of the ocean. The situation in the icecovered areas in the north remains to be investigated.
The Barents Sea is a spring bloom system. During winter, primary production is close to zero. Timing of the phytoplankton bloom varies throughout the Barents Sea and there may also be a high inter-annual variability. The spring bloom starts in the south-western areas and spreads north and east with the retracting ice. In early spring, the water is mixed from surface to bottom. Despite adequate nutrient and light conditions for production, the main bloom does not occur until the water becomes stratified.

Stratification of water masses in different areas of the Barents Sea may occur in several different ways; 1) through fresh surface water from melting ice along the marginal ice zone; 2) through solar heating of surface layers in Atlantic water masses; or 3) through lateral dispersion of waters in the southern coastal region (Rey, 1981). As in other areas, diatoms are also the dominant phytoplankton groups in the Barents Sea (Rey, 1993). Diatoms particularly dominate the first part of the spring bloom, and the concentration of diatoms can reach up to several million cells per litre. They require silicate for growing, and when this is consumed, other phytoplankton groups, such as flagellates, take over. An important flagellate species in the Barents Sea is Phaeocystis pouchetii but other species may, however, predominate the spring bloom in different years.

## Zooplankton

In the Barents Sea ecosystem, zooplankton forms a link between phytoplankton (primary producers) and fish, mammals and other organisms at higher trophic levels. Zooplankton biomass in the Barents Sea can vary significantly between years and crustaceans are important. The calanoid copepods of the genus Calanus play a key role in this ecosystem. Calanus finmarchicus, is most abundant in Atlantic waters and C. glacialis is most abundant in Arctic waters. Both form the largest component of zooplankton biomass.

Calanoid copepods are largely herbivorous, and feed particularly on diatoms (Mauchline, 1998). Krill (euphausiids), another group of crustaceans, also play a significant role in the Barents Sea ecosystem as food for fish, seabirds, and marine mammals. Krill species are believed to be omnivorous: filter-feeding on phytoplankton during the spring bloom; while feeding on small zooplankton during other times of the year (Melle et al., 2004). Four dominant species that occupy different niches in the community of Barents Sea euphausiids are: Meganyctiphanes norvegica (neritic shelf boreal); Thysanoessa longicaudata (oceanic arcto-boreal); T. inermis (neritic shelf arcto-boreal); and T. raschii (neritic coastal arcto-boreal) (Drobysheva, 1994). The two latter species comprise $80-98 \%$ of total euphausiid abundance, but species composition may vary between years relative to climate (Drobysheva, 1994). After periods with cold climate, observed abundance of $T$. raschii increased while abundance of $T$. inermis decreased (Drobysheva, 1967). Advection from the Norwegian Sea is influenced by the intensity of Atlantic water inflow, which also influences the composition of species (Drobysheva, 1967; Drobysheva etal., 2003).

Three amphipod species were found abundant in the Barents Sea; Themisto abyssorum and T. libellula in the western and central Barents Sea, and T. compressa is found, albeit less abundant, in central and northern regions. T. abyssorum is most abundant in subArctic waters. In contrast, the largest of the Themisto species, T. libellula, is largely restricted to combined Atlantic and Arctic water masses. High abundance of T. libellula was observed adjacent to the Polar Front. Amphipods feed on small zooplankton and copepods form an important component of their diet (Melle et al., 2004).
"Gelatinous zooplankton" is a term often used by non-specialists in reference to classes of organism that are jelly-like in appearance. The term "jelly fish" is commonly used in reference to marine invertebrates belonging to the class Scyphozoa, phylum Cnidaria. Neither of these terms implies any systematic relationship to vertebrate fish. The term "jelly fish" is also often used in reference to relatives of true scyphozoans, particularly the Hydrozoa and the Cubozoa. Both comb-jellies (Ctenophora) and "true" jelly fish are predators, and they compete with plankton-eating fish, because copepods often are significant prey items.

## Benthos

The sea floor is inhabited by a wide range of organisms. Some are buried in sediment, others are attached to a substrate, some are slow and sluggish, others roving and rapid. Many feed by actively or passively, sieving food particles or small organisms from the water. Others eat the bottom sediments (detritus feeders), eat carrion (scavengers) or hunt other animals (carnivores). The high diversity among bottom animals is presumed to be due to the abundance of micro-habitats that organisms can adapt. In shallow waters, kelp forests are feeding and nursery habitats for many species of fish, birds, and mammals. Below the sublittoral zone, sea anemones, sponges, hy drozoans, tunicates, echinoderms, crustaceans, molluscs and many other animal groups
abound on hard substrates. These large conspicuous animals are not abundant on sand or muddy bottoms, and in fact some of these habitats may at first look rather lifeless. However, most of the benthic animals in these habitats live buried in the sediments. Polychaete worms, crustaceans and bivalves are found in the sediments well as a myriad of other taxa. Some muddy areas might have dense aggregations of brittle stars, sea stars or bivalves.

More than 3050 species of benthic invertebrates inhabit the Barents Sea (Sirenko, 2001). The benthic ecosystems in the Barents Sea have considerable value, both in direct economic terms, and in their ecosystem functions. Scallops, shrimp, king crab, and snow crab arebenthic residents which areharvested in the region. Many species of benthos are also interesting for bio-prospecting or as a future food resource, such as sea cucumber, snails and bivalves. Several of them are crucial to the ecosystem. Important fish species such as haddock, catfish and most flatfishes primarily feed on benthos. Many benthic animals, primarily bivalves, filter particles from the ocean and effectively clean it up. Others scavenge on dead organisms, returning valuable nutrients to the water column. Detritus feeders and other active diggers regularly move the bottom sediments around and therefore increase sediment oxygen content and overall productivity - much like earthworms on land.

The northern shrimp (Pandalus borealis) is distributed in most deep areas of the Barents Sea and Spitsbergen waters. The densest concentrations are found in depths between 200 and 350 meter. The shrimp mainly feed on detritus, but may also be a scavenger. Shrimp is also important as a food item for many fish species and seals.
Red king crab (Paralithodes camtschatica) was introduced to the Barents Sea in the 1960s. Presently it is an important commercial species. Adult red king crabs are opportunistic omnivores.
The snow crab (Chionoecetes opilio) is an invasive species in the Barents Sea. The first recordings of this species in the Barents Sea were in 1996. Since 2003 snow crab have been found in the stomachs of cod, haddock, catfishes and thorny skates that indicates that the crab abundance and settlement density substantially increased.

## Fish

More than 200 fish species are registered in trawl catches during surveys of the Barents Sea, and nearly 100 of them occur regularly. Even so, the Barents Sea is a relatively simple ecosystem, with few fish species of potentially high abundance. Different species of fish are not evenly distributed throughout the Barents Sea. Rather, they exhibit highest abundance in areas with suitable environmental conditions. Commercially important fish species include Northeast Arctic cod, Northeast Arctic haddock, Barents Sea capelin, polar cod and immature Norwegian springspawning herring. In warm years, increased numbers of young blue whiting have migrated into the Barents Sea. Species distribution largely depends on positioning of the Polar Front. There have been significant variations in abundance of these species. These variations are due to a combination of fishing pressure and environmental variability Cod, capelin, and herring are key species in the Barents Sea trophic system.

In general the four pelagic species (herring, capelin, polar cod and blue whiting) have minor overlapping distributions; with the blue whiting in the west, the herring in the south, the polar cod in the east (except for an overlapping part of the stock in the Svalbard region) and the capelin in the north and central areas. In southwestern areas blue whiting and herring partly overlap. However, they occupy different parts of the water column.

The recruitment of the Barents Sea fish species has shown a large year-to-year variability (Tables 1.1-1.2). The most important reasons for this variability are variations in the spawning biomass, climate conditions, food availability and predator abundance and distribution. Variation in the recruitment of some species, like cod, haddock and herring, has been associated with changes in the influx of Atlantic waters into the Barents Sea.

Cod prey on capelin, herring, and smaller cod; while herring prey on capelin larvae. Cod is the most important predator fish species in the Barents Sea, and feeds on a wide range of prey, including larger zooplankton, most available fish species and shrimp (Table 1.3-

Table 1.6). Cod prefer capelin as a prey, and fluctuations of the capelin stock (Table 1.7) have a strong effect on growth, maturation and fecundity of cod, as well as on cod recruitment because of cannibalism. The role of euphausiids for cod feeding increases in the years when capelin stock is at a low level (Ponomarenko and Yaragina $1990)$. Also, according to Ponomarenko $(1973,1984)$ inter annual changes of euphausiid abundance is important for the survival rate of cod during the first year of life.

Capelin feed on zooplankton produced near the ice edge. Farther south, capelin is the most important prey species in the Barents Sea as it transports biomass from northern to southern regions (von Quillfeldt and Dommasnes, 2005). The capelin has showed large variations in abundance.

Herring is also a major predator on zooplankton. The herring spawns along the Norwegian western coast and the larvae drifts into the Barents Sea. The juveniles of the Norw egian spring-spawning herring stock are distributed in the southern parts of the Barents Sea. They stay in this area for about three years before they migrate west and southwards along the Norw egian coast and mix with the adult part of the stock. The presence of young herring in the area has a profound effect on the recruitment of capelin, and it has been shown that when rich year classes of herring enters to the Barents Sea, the recruitment to the capelin stock is poor, and in the following years the capelin stock collapses (Gjøsæter and Bogstad, 1998).
Haddock is also a common species, and migrates partly out of the Barents Sea. The stock has large natural variations in stock size. Food composition of haddock consists mainly ofbenthic organisms.
Saithe is found mainly along the Norwegian coast, but also occurs in the Norwegian Sea and in the southern Barents Sea. The 0 -group saithe drifts from the spawning grounds to inshore waters. The smaller individuals feed on crustaceans, while larger saithe depends more on fish as prey (Dolgov, WD 29, AFWG 2006; Mehl, WD7, AFWG 2005). The main fish prey is young herring, Norway pout, haddock, blue whiting and capelin, while the dominating crustacean prey is krill. Polar cod is a cold-water species found particularly in the eastern Barents Sea and in the north. It seems to be an important forage fish for several marine mammals, but to some extent also for cod. There is little fishing on this stock.

Deep-sea redfish and golden redfish used to be important elements in the fish fauna in the Barents Sea, but due to heavy over fishing these stocks declined strongly during the 1980's, and has since then stayed at a low level. Young redfish are plankton eaters, but larger individuals take larger prey, including fish.

Greenland halibut is a large and voracious fish predator with the continental slope between the Barents Sea and the Norw egian Sea as its most important area, but it is also found in the deeper parts of the Barents Sea. Investigations in the period 19801990 show ed that cephalopods (squids, octopuses) dominated in the Greenland halibut stomachs, as well as fish, mainly capelin and herring. Ontogenetic shift in prey preference was clear with decreasing proportion of small prey (shrimps and small capelin) and increasing proportion of larger fish with increasing predator length. The largest Greenland halibut (length more than $65-70 \mathrm{~cm}$ ) had a rather big portion of cod and haddock in the diet.

The blue whiting has its main distribution area in the Norwegian Sea and Northeast Atlantic, and the marginal northern distribution is at the entrance to the Barents Sea. Usually the blue whiting population in the Barents Sea is small. In years with warm Atlantic water masses the blue whiting may enter the Barents Sea in large numbers, and the blue whiting can be a dominant species in the western areas. This situation occurred from 2001 onwards, and blue whiting were found in great numbers for the period 2003-2007. Since then it has decreased strongly again. This rise and fall is probably due to a combination of variation in stock size and environmental conditions. In the diet of blue whiting zooplankton(copepods, hyperiids and euphausiids) is dominant in the younger age groups, while fish is increasingly important as the blue whiting gets older(Dolgov, WD 29, AFWG 2006).

Long rough dab is a typical ichthyobenthophage, which mainly eats benthos (ophiura, polychaetes etc.) and different fish species (Dolgov, WD 29, AFWG 2006). At older stages the proportion of fish in the diet increases (polar cod and cod, capelin and juvenile redfish). The larger long rough dab also feed on their own juveniles and juvenilehaddock.

Thorny skate preys primarily on large crustaceans, shrimps and crabs (Dolgov, WD 29, AFWG 2006), but may also in a lesser extent feed on fish. The most common fish species are young cod and capelin. Round skate fed mainly on benthos, especially Polychaeta and Gammaridae. Arctic skate feed mainly on fish and shrimp (herring, capelin, redfish and northern shrimp). Blue skate diet consists largely of fish, mainly young cod and haddock, redfish, and long rough dab). Spinytail skate also prey mostly on fish, which included haddock, redfish and long rough dab. Total yearly food consumption by thorny skate is estimated to be around 160 thousand tonnes, of which around 75 thousand tonnes comprised commercial fishes and invertebrates. Total yearly food consumption by all other skate species was estimated to be around 30 thousand tonnes, of which around 20 thousand tonnes was commercial species (Dolgov,WD 29, AFWG 2006).

## Mammals and seabirds

Marine mammals, as top predators, are keystone species significant components of the Barents Sea ecosystem. About 25 species of marine mammals regularly occur in the Barents Sea, including: 7 pinnipeds (seals and walruses); 12 large cetaceans (large whales); 5 small cetaceans (porpoises and dolphins); and the polar bear (Ursus maritimus). Some of these species are not full-time residents in the Barents Sea, and use temperate areas for mating, calving, and feeding (e.g. minke whale Balaenoptera acutorostrata). Others reside in the Barents Sea all year round (e.g. white-beaked dolphin Lagenorhynchus albirostris, and harbour porpoise Phocoena phocoena). Some marine mammals are naturally rare, such as the beluga whale Delphinapterus leucas. Others are rare
due to historic high exploitation, such as bowhead whale Balaena mysticetus and blue whale Balaenoptera musculus.

Marine mammals may consume up to 1.5 times the amount of fish caught in fisheries. Minke whales and harp seals may each year consume 1.8 million and 3-5 million tons of prey of crustaceans, capelin, herring, polar cod, and gadoid fish respectively (Folkow et al., 2000; Nilssen et al., 2000). Functional relationships between marine mammals and their prey seem closely related to fluctuations in marine ecosystems. Both minke whales and harp seals are thought to switch between krill, capelin and herring depending on availability of the different prey species (Lindstrøm et al., 1998; Haug et al., 1995; Nilssen et al., 2000).

Fish and mammals have seasonal feeding migrations so that the stocks in the area will have their most northern and eastern distribution in August-September and be concentrated in the southern and south-w estern areas in February-March. The Barents Sea has one of the largest concentrations of seabirds in the w orld (Norderhaug et al., 1977; Anker-Nilssen et al., 2000); its 20 million seabirds harvest annually approximately 1.2 million tonnes of biomass from the area (Barrett et al., 2002). Nearly 40 species are thought to breed regularly in northern regions of the Norwegian Sea and the Barents Sea. Abundant species belong to the auk and gull families. Seabirds play an important role in transporting organic matter and nutrients from the sea to the land (Ellis, 2005). This transport is of great importance especially in the Arctic, where lack of nutrients is an important limiting factor.

## Rare, threatened and invasive species and infectious organisms

There are 10 types of parasites found in the fish of the Barents Sea, but it is hard to determine which groups of parasitic organisms that play an important role in the population dynamics of their hosts. The Barents Sea parasites considered to be most damaging to the human health are larvae stages of Cestoda (Diphyllobothrium and Pyramicocephalus genera), Nematoda (Anisakis and Pseudoterranova genera) and Palaeacanthocephala (Corynosoma genera). 82 species of helminthes are recorded from 18 bird species. The Barents Sea birds' helminthofauna mostly consists of the species with the life cycle dependent on coastal ecosystems. Invertebrates and fish from the littoral and upper sub littoral complex serve as their intermediate hosts.

The Barents Sea includes species that either have very small populations or species that have recently undergone considerable population decline (or are expected to do so in the close future). The assessments are done by use of the IUCN criteria (IUCN, 2001; 2003), but the Global, the Russian and the Norwegian lists available cannot be directly compared. All these lists are closely related and have high relevance for the conservation of biodiversity, and the list from the Barents Sea include a total of 56 species comprising of 28 fish species, 9 bird species, and 18 mammal species.

Invasions of alien species - spread of the representatives of various groups of living organisms beyond their primary habitats - are global in nature. Their introduction and further spread often leads to the undesirable environmental, economic and social consequences. Different modes of biological invasions can be natural movement associated with the population dynamics and climatic changes, intentional introduction and reintroduction, and accidental introduction with the ballast waters and along with the intentionally introduced species, etc. The best known examples of introduced species in the Barents Sea are red king crab (Paralithodes camtschaticus) and snow crab (Chionoecetes opilio).

## Human activity

The Barents Sea is strongly influenced by human activity; historically involving the fishing and hunting of marine mammals. More recently, human activities also involve transportation of goods, oil and gas, tourism and aquaculture. In the last years interest has increases on the evaluation of the most likely response of the Barents Sea ecosystem to the future climate changes due to anthropogenic effects on climate warming.

Fishing is the largest human impact to the fish stocks in the Barents Sea, and thereby the functioning of the whole ecosystem. However, the observed variation in both fish species and ecosystem is also impacted by other effects such as climate and predation. The most widespread gear used in the central Barents Sea is bottom trawl, but also long line and gillnets are used in the demersal fisheries. The pelagic fisheries use purse seine and pelagic trawl.

The Barents Sea remains relatively clean, however, when compared to marine areas in many industrialized parts of the world. Major sources of contaminants in the Barents Sea are natural processes, long-range transport, accidental releases from local activities, and ship fuel emissions. Results of recent studies indicate low level of contaminants in the Barents Sea marine environment and confirm results of earlier studies on bottom sediments in the same areas. In the near-term, observed levels of contaminants in the marine environment should not have significant impact on commercially important stocks and on the Barents Sea ecosystem as a whole.

Traditionally, fishing having been the most important and far-reaching human activity in the ecosystem has been given most of the attention with analyses of impacts and risks. This need has increased in importance as oil- and gas industries have begun to develop new off-shore fields in the Barents Sea, and ship transport of oil and gas from the region has increased exponentially over the last 5 years.
The Barents Sea can become an important region for oil and gas development. Currently offshore development is limited both in the Russian and Norwegian economic zones (to the Snøhvit field north of Hammerfest in the Norwegian zone), but this may increase in the future with development of new oil- and gas fields. In Russia there are plans for the development of Stockman, a large gas-field west of Novaya Zemlja. The environmental risk of oil and gas development in the region has been evaluated several times, and is a key environmental question facing the region.

Transport of oil and other petroleum products from ports and terminals in NWRussia have been increasing over the last decade. In 2002, about 4 million tons of Russian oil was exported along the Norwegian coastline, in 2004, the volume reached almost 12 million tons, but the year after it dropped, and from 2005 to 2008 was on the levels between 9,5 and 11,5 million tons per year. In a five-ten years perspective, the total available capacity from Russian arctic oil export terminals can reach the level of 100 million tons/year (Bambulyak and Frantsen, 2009). Therefore, the risk of large accidents with oil tankers will increase in the years to come, unless considerable measures are imposed to reduce such risk.

Tourism is one of the largest and steadily growing economic sectors world-wide. Travels to the far north have increased considerable during the last 15 years, and there are currently nearly one million tour ists annually.

The high biodiversity of the oceans represents a correspondingly rich source of chemical diversity, and there is a growing scientific and commercial interest in the
biotechnology potential of Arctic biodiversity. Researchers from several nations are currently engaged in research that could be characterised as bio-prospecting.

Aquaculture is growing along the coasts of northern Norway and Russia, and there are several commercial fish farms producing salmonids (salmon, trout), white fish (mainly cod) and shellfish.

Ocean acidification is greater and happening faster than any previous acidification process experienced in millions of years. The absorption of $\mathrm{CO}_{2}$ generally goes faster in colder waters and thus will rapidly affect the Barents Sea.

### 1.2 State and expected situation of the ecosystem (Figures 1.2-1.10, Tables 1.3-1.6, 1.9)

### 1.2.1 Climate

## Atmospherical conditions

In 2009, the weather over the North Atlantic was determined by cyclonic activity throughout the year, and northerly and easterly winds prevailed over the Barents and northern Norwegian Seas. In winter, spring and autumn, air temperature averaged over the western and eastern parts of the Barents Sea was higher than normal, with maximum positive anomalies $\left(3.9-4.1^{\circ} \mathrm{C}\right)$ in the eastern Barents Sea in January and March. In summer, positive anomalies did not exceed $1^{\circ} \mathrm{C}$, and small negative anomalies were observed in some months

## Water temperature

In general the temperatures in the entire Barents Sea in 2009 was still high (about 0.5$1.0^{\circ} \mathrm{C}$ above the long-term average), and at about the same levels as in 2008. At the end of the year the temperature in the Atlantic water masses was increasing again. In the beginning of 2010 the temperature decreased again, but is still above the longterm mean.

Sea surface temperature (SST) in the Barents Sea showed much of the same variations as the air temperatures. In winter, due to the warmer-than-usual air masses over the central and eastern Barents Sea and therefore the less-than-usual atmospheric cooling, the SST was higher than normal, with maximum positive anomalies $\left(1.0^{\circ} \mathrm{C}\right)$ in the central part of the sea. In the western and north-western Barents Sea, on the contrary, the SST was lower than normal throughout most of the year, with maximum negative anomalies $\left(-0.5^{\circ} \mathrm{C}\right)$ in April and July. The weaker-than-usual spring-and-summer warming caused decreasing SST anomalies. From June to August, negative anomalies of SST were observed in most of the sea. In autumn, SST anomalies increased due to the intensification of cyclonic activity and warm air-masses transport; maximum positive anomalies of SST (up to $1.6^{\circ} \mathrm{C}$ ) were found in the southern areas in November.

Development in the coastal waters is measured at the Ingøy fixed station, and show that during 2009 the surface temperature was only slightly above normal through most of the year except in late fall/early winter 2009/2010. In the deeper waters (at 250 m ), which is strongly influenced by Atlantic Water, the temperature was above normal throughout the year. In both the surface and deeper layers, the temperature increased (relative to the normal) in late fall 2009/early winter 2010, but decreased again in spring 2010, with surface temperatures around and deeper layers still slightly above the long term mean.

The Fugleya-Bear Island and Vardø-North sections, which capture all the Atlantic Water entering the Barents Sea from south-west, showed temperatures close to $0.5^{\circ} \mathrm{C}$ above the long-term mean in early 2009 (Figure 12). This is lower than the last 5-6 winter, and is due to low er air temperatures causing more intense heat loss in combination with weak inflow of Atlantic Water. Over the year the temperatures increased, and in October 2009 the temperature in south-west was $0.9{ }^{\circ} \mathrm{C}$ above the long-term mean. The annual mean temperature in 2009 was close to the year of 2008. In the beginning the temperature at the Vardø-North decreased again to $\sim 0.5{ }^{\circ} \mathrm{C}$ above the long term mean.
Temperature in the upper 200 m layer in the southern Barents Sea (Kola section) was higher than normal throughout the year of 2009, and, during the second half of the year, it was higher than in 2008 (Figure 1.3). At the beginning of the year, the weaker-than-usual seasonal cooling caused an increase in positive temperature anomalies (by $0.1-0.3^{\circ} \mathrm{C}$ ) in the Atlantic water compared to December of 2008. The positive anomalies changed slightly during the first half of the year, then they decreased to September due to easterly and northeasterly winds prevailed in spring and summer. During autumn, temperature anomalies in the main warm currents increased again due to the intensification of cyclonic activity and air-mass transport from the west. By December, temperature anomalies exceeded $1.0^{\circ} \mathrm{C}$ in all parts of the Kola Section, and the highest December temperature for the period from 1951 to the present was observed in the Murman Current. The annual temperature in the Murman Current in 2009 was ty pical of anomalous warm years and close to that of 2008.
Temperature in the bottom layer of the Barents Sea in August-September 2009 was typical of warm and anomalous warm years. Positive temperature anomalies were observed in most of the surveyed area and were, on average, $0.3-1.0^{\circ} \mathrm{C}$. The largest positive temperature anomalies $\left(>1.5^{\circ} \mathrm{C}\right)$ were observed in the eastern Barents Sea, in the areas adjacent to the Eastern Basin (Figure1.4). Compared to 2008, the volume of cold Arctic waters increased significantly in the northern Barents Sea, and for the first time in the last three years waters with negative temperature were found in the Eastern Basin. So, in comparison with the previous year, it caused decrease in the spatially averaged bottom temperature of the surveyed area except the southern Barents Sea occupied by the Murman Current and the Central branch of the North Cape Current. In the beginning of 2010 the bottom temperatures in the south and southwestern parts were higher than in the same period in 2009, while they were lower in the deep central parts.

According to computations with a prediction model, based on harmonic analysis of the Kola Section temperature time series, the temperature of the Atlantic water in the Murman Current in 2010-2011 is expected to decrease to values typical of warm years, namely to $4.5 \pm 0.5^{\circ} \mathrm{C}$ (with anomaly of $+0.6^{\circ} \mathrm{C}$ ) in 2010 and to $4.4 \pm 0.5^{\circ} \mathrm{C}$ (with anomaly of $+0.5^{\circ} \mathrm{C}$ ) in 2011. The years of 2010 and 2011 are similar to 1989,1991, 2001 and 2002.

## Salinity

The salinity variations show a close resemblance to temperature, although not completely. In Fugleya-Bear Island the salinity has been decreasing since 2006, while in Vardø-N it has increased over the last years. Salinity in the Atlantic water masses in 2009 was still high compared to the long term trend.

## Inflow of Atlantic water

The volume flux of Atlantic Water flowing into the Barents Sea is predominantly barotropic, with large fluctuations in both current speed and lateral structure. In general, the current is wide and slow during summer and fast, with possibly several cores, during winter. The mean transport of Atlantic Water into the Barents Sea for the period 1997-2009 is $2 \mathrm{~Sv}\left(\mathrm{~Sv}=10^{6} \mathrm{~m}^{3} \mathrm{~s}^{-1}\right)$ with an average of 2.2 Sv during winter and 1.8 Sv during summer. During years in which the Barents Sea changes from cold to warm marine climate, the seasonal cycle can be inverted. Moreover, an annual event of northerly wind causes a pronounced spring minimum inflow to the western Barents Sea; at times even an outward flow.

The time series of volume transport reveals fluxes with strong variability on time scales ranging from one to several months (Figure 1.5). The strongest fluctuations, especially in the inflow, occur in late winter and early spring, with both maximum and minimum in this period. The recirculation seems to be more stable at a value of something near 1 Sv , but with interruptions of high outflow episodes.

The volume flux varies with periods of several years, and was significantly lower during 1997-2002 than during 2003-2006. The year of 2006 was a special year as the volume flux both had a maximum (in winter 2006) and minimum (in fall 2006). Since then the inflow has been low, particularly during spring and summer. The inflow in 2009 was much as in 2007 and 2008; moderate during winter followed by a strong decrease in spring. In early summer 2009 the flux was close to 1.5 Sv below the average. As the observational series still only have data until summer 2009, it cannot give information about the situation in fall 2009 and early winter 2010. There is no significant trend in the observed volume flux from 1997 to summer 2010.

## Ice conditions

The variability in the ice coverage in the Barents Sea is linked to the temperature of the inflowing Atlantic water, the northerly winds, and import of ice from the Arctic Ocean and the Kara Sea. The ice has a response time on temperature changes in the Atlantic inflow (onetwo years), but usually the sea ice distribution in the western Barents Sea respond a bit quicker than in the eastern part. Due to the high temperatures there has been little ice in the last years (Figure 1.6). During the period 20032006 the winter ice edge had a substantial retreat towards north-east, but since then the ice area has increased.

For the first eight months of the year of 2009, the sea ice extent in the Barents Sea was less than normal, but more than in 2008. In comparison with the previous year, the ice coverage (expressed as a percentage of the sea area) was $10-18 \%$ more in JanuaryMay, $5-9 \%$ more in June and August and the same in July. Ice melting in summertime was more intensive than in 2008. By July, the south-eastern Barents Sea was ice free, which is almost one month earlier than in 2008. Ice formation started in the northernmost sea only at the end of October. In October, the ice coverage was $13 \%$ less than normal and $5 \%$ less than in 2008. By December, the ice coverage of the Barents Sea was still lower than normal but higher than in 2008, a situation that continued into the beginning of 2010.
It is expected that there will be slightly less or around average ice conditions in 2010.

## Hydrochemical conditions

According to the chemical observations along the Kola Section in 2009, some decrease in oxygen saturation of the bottom layer was found in the southern Barents Sea compared to 2008: the oxygen saturation anomaly averaged from January to October was $-0.24 \%$ in 2009, and $0.78 \%$ in 2008. Negative anomalies prevailed at the beginning of the year, while small positive anomalies prevailed in summer and autumn.

### 1.2.2 Phytoplankton

In Norwegian waters there was not observed any large aberration in the annual succession in the phytoplankton along the fixed transect (Vardø - North and FugloyaBear Island) in 2009. The spring bloom occurred from mid March to mid April within the "normal" period of the spring bloom at the Bear Island transect. The bloom starts in the coastal waters "spreading" out into the open areas. In April the diatoms were dominating. During summer the phytoplankton was compound of small flagellates, dinoflagellates, and at some stations diatoms. During autumn larger dinoflagellates was common, however, at some stations diatoms had moderate to high abundance.

### 1.2.3 Zooplankton

The mesoplankton biomass measured in August-September 2009 was clearly below the long-term mean in the Norwegian sector but with slightly higher values along the border to the Russian zone. A particular feature in 2009 is the very high biomass found in the Russian sector north of $75^{\circ} \mathrm{N}$ and east of $40^{\circ} \mathrm{E}$. The average zooplankton biomass in the western and central Barents Sea in 2009 was 5.87 g dry weight $\mathrm{m}^{-2}$ compared to 6.48 g in 2008 and 7.13 g in 2007 (Figure 1.7).
The macroplankton survey conducted in autumn and winter 2009 showed that on average, abundance of euphausiids in the west and northwest of the sea was close to the level of 2008 (Figure 1.8). However, in the center, east and coast areas the abundance indices of krill increased 1.5-2 times compared to 2008. In total the macroplankton survey showed that the abundance indices of euphausiids were above than the long-term mean.
The average zooplankton abundance in 2009, together with the considerable decline observed since 2006, suggest that the condition for local production is less favourable for 2010. The total production will probably depend largely on the magnitude of zooplankton advection from the Norwegian Sea. The macroplankton feeding conditions for planktivorous fish in 2010 is expected to be similar to 2009.
The abundance of gelatinous zooplankton, caught by pelagic trawling, show a lower abundance of gelatinous zooplankton in 2009 compared to 2008. Both in 2008 and in 2009, the distribution of "jelly fish" also showed a considerable overlap with regions poor in mesozooplankton biomass.

### 1.2.4 Northern shrimp

According to the Russian-Norwegian ecosystem survey in August - September 2009 the largest catches of the northern shrimp were recorded in the eastern and northern Barents Sea and north of Spitsbergen. The investigations of 2009 showed that the total stock of the northern shrimp increased compared to last year.

### 1.2.5 Fish

The current and expected situation of the commercial stocks in the Barents Sea addressed by the AFWG is given in later chapters. Therefore focus in this subchapter is on other main species that interacts with the AFWG stocks, and on the role of the AFWG species in an ecosystem perspective (e.g. as predators). Special attention is given when there are deviations from the general situation. An overview of the development of pelagic and demersal stocks is given in Figures 1.9 and 1.10.

## NEA cod consumption

The food consumption of cod in 1984-2009, based on data from the Joint RussianNorwegian stomach content data base, is presented in Table 1.3-1.4. The main prey items in 2009 were capelin, polar cod, krill, haddock, herring, shrimp, cod and amphipods. In comparison with 2008 the importance of capelin and herring has increased while the importance of krill and shrimp has decreased. The consumption calculations madeby IMR show that the total consumption by age 1 and older cod in 2009 was about 6 million tonnes (Table 1.3), while similar calculations by PINRO gave about 5 million tonnes. According to calculations by IMR and PINRO the consumption per cod was about the same in 2009 as in 2008 (Tables 1.5-1.6).

## Blue whiting and polar cod

Based on the most recent estimates of fishing mortality and SSB, ICES classifies the stock as having full reproductive capacity, and being harvested sustainably. SSB increased to a historical high in 2004 but has decreased since, and is expected to be just above $\mathrm{B}_{\mathrm{pa}}$ in 2011. The estimated fishing mortality is slightly below $\mathrm{F}_{\mathrm{pa}}$. Recruitment in 1995-2004 was at a much higher level than earlier, but the 2005 and later year classes seem to be poor. Total landings in 2008 were 1.3 mill. tonnes, which is low er than in 2007. Blue whiting is not fished in the Barents Sea.

The high abundance of blue whiting in the Barents Sea in 2004-2007 may be due to a large stock size in this period combined with high temperature. Blue whiting has been observed in the western and southern Barents Sea for many years, but never in such quantities, and never as far east and north in this area as in 2004-2007. In autumn 2009, the acoustic abundance of blue whiting was estimated to 0.3 million tonnes, which is higher than in 2008, but still low. Also, the swept area estimate of blue whiting in winter 2010 was the lowest in the time series, which go back to 2001. Thus, the abundance of blue whiting in the Barents Sea is expected to stay at a low level until the recruitment to the stock increases again.

The polar cod stock is presently at a high level. Norway took some catches of polar cod in the 1970s and Russia has fished on this stock more or less on a regular basis since 1970 . The stock size has been measured acoustically since 1986 and the stock has fluctuated between 0.1-1.9 million t . In 2009, the stock size was measured to about 0.9 million $t$., which is below the estimate obtained in 2008. The natural mortality rate in this stock seems to be very high, and this is explained by the importance of polar cod as prey for cod and different stocks of seals.

## Herring and capelin

Based on the most recent estimates of SSB and fishing mortality, ICES classifies the stock as having full reproductive capacity and being harvested sustainably. The 1998, 1999, 2002 and 2004 year classes dominate the current spawning stock which is esti-
mated to 12.2 million $t$ in 2010. Preliminary indications show that the year classes 2005-2009 are below average. Therefore the abundance of herring in the Barents Sea is believed to be at a relatively low level in 2010. This stock has shown a large dependency on the occasional appearance of very strong year classes. In recent years the stock has tended to produce strong year classes more regularly. However, as no strong year classes have been produced since 2004, the stock is expected to decline. Norwegian spring-spawning herring is fished along the Norwegian coast and in the Norwegian Sea, but not in the Barents Sea. How ever, juveniles from this stock play an important part role in the ecosystem in the Barents Sea.

The capelin stock size is at a level somewhat above average. Based on the most recent estimates of SSB and recruitment ICES classifies the stock as having full reproductive capacity. The maturing component in autumn 2009 was estimated to be 2.3 mill t ., and SSB 1st April 2010 was predicted to be at 0.52 mill t. The spawning stock in 2010 consisted of fish from the 2006 and 2007 year classes, but the 2006 year class dominated. The survey estimate at age 1 of the 2008 year class is somewhat below the long-term average. Observations during the international 0-group survey in AugustSeptember 2009 indicated that the 2009 year class is below average.

The estimated annual consumption of capelin by cod has varied between 0.2 and 3.0 million $t$ over the period 1984-2009. Young herring consume capelin larvae, and this predation pressure is thought to be one of the causes for the poor year classes of capelin in the periods 1984-1986, in 1992-1994, and from 2002-2005.

## Non-commercial species

Thorny skate (Amblyraja radiata) was quite widely distributed in the Barents Sea except for the south eastern and north eastern regions, as in 2008. The observed abundance of this species was higher than in 2008. The thorny skate preferred to keep in a wide range of depths from 50 down to 300 meters.

Northern skate (Amblyraja hyperborea) was distributed in the northeast part of the Barents Sea and along the shelf slope to the west of Spitsbergen. It was mainly found in the depth range 200 to 300 meters.

Plaice (Pleuronectes platessa) $w$ as distributed in a range of depths from 50 down to 100 meters) on northwest from Kanin peninsula.

According to observations in 2009 the tendency of expansion of Norway pout (Trisopterus esmarkii) in the Barents Sea is continuing. Main density concentrations of Norway pout were registered in the south-western areas. At the same time, along the warm Spitsbergen current, Norway pout was observed until $81^{\circ} \mathrm{N}$. Along coastal North Cape current Norway pout were distributed eastward up to $47^{\circ}$ E. It seems like Norway pout have occupied the blue whiting distribution area after this species declined.

In the ecosystem survey in 2009 there were both new species to the area and recordings of rare species in the area of observation. Some of these species have their main distribution in the warm waters of the Norwegian Sea (Molva molva, Schedophilus medusophagus) or in the cold waters of the Kara Sea (Arctogadus glacialis) bordering the Barents Sea.

### 1.2.6 Marine mammals

## Harp Seal

Since 1998 the abundance of harp seal pup production in the White Sea has been sharply reduced, according to the PINRO aerial surveys. How ever the decrease in the harp seal pup production abundance has become slower recently and even some slight increase has been observed. The abundance of harp seal pups in the whelping patches in 2009 calculated using the data from aerial surveys was more than two times lower, compared to the data obtained for 2000-2003.

One of the key factors, which caused the reduction in the harp seal pup abundance in 2004-2009, was the diminished ice extent due to warming. The changed ice conditions were responsible for the redistribution of animals in the pup period. Abnormal ice conditions in the White Sea in 2005-2009 possibly also led to higher natural mortality of pups.

The decrease in the abundance of harp seal pup production leads to a reduction of the whole harp seal population (the model estimate for 2009-1.2 million animals).

## Predation by mammals

Analyses of consumptions by marine mammals in the Barents Sea for 2009 are not available. Last estimates are shown in Table 1.9.

### 1.2.7 Future long-term trends

This section is a short version of Stiansen et al (2009).
Air temperatures have increased almost twice as fast in the Arctic than the global average over the last 50 years. Models predict that air temperatures will continue to increase considerably. With the accelerated increase in air temperatures it is predicted that summer sea ice will disappear. Polar Front that separates the cold Arctic and warm Atlantic waters will move farther north and east. Although long-term climate projections are associated with considerable uncertainty, it is highly likely, however, that any significant warming will cause shifts in species ranges and changes in their production. The expected northward extension of warm Atlantic water will lead in general to that temperate zooplankton would shift northward while ice fauna, such as the large amphipods would diminish due to a massive loss of habitat because of the disappearance of multi-year ice (Skjoldal et al., 1987; Loeng et al., 2005). Ellingsen et al. (2008) also predicted that the Atlantic zooplankton production, primarily Calanus finmarchicus, would increase by about $20 \%$ and spread farther eastward while the Arctic zooplankton biomass would decrease significantly (by $50 \%$ ) resulting in an overall decrease in zooplankton production in the Barents Sea.
A number of fish species, e.g. cod and capelin, will likely have a more northern and/or eastern distribution and boreal species such as blue whiting and mackerel may become common in the Barents Sea. These changes will likely result in potentially large changes in community composition and it is possible that the structure of the ecosystem may shift irreversibly. In addition, sea ice extent will be reduced, and this will have a negative impact on ice-dependent flora and fauna, such as polar bears. Reduction in sea ice extent may also lead to increased primary productivity, if nutrient supply is not reduced significantly due to increased stratification in the water column. An increase in primary productivity coupled with other positive effects of increased temperature on fish growth and reproduction, may cause productivity of
cod, haddock and other commercially important species to increase. How ever, negative effects on prey species may also occur. Thus, overall effects on fish productivity are hard to predict.
Similarly, the many complex ways in which species interact creates considerable uncertainty in any set of predictions as to what the overall response of climate warming to the ecosystem will be. If warming causes phytoplankton to increase, this is expected to result in an overall increase in fish production. For example, model studies show that higher primary production tends to lead to an increase in cod recruitment in the Barents Sea (Svendsen et al., 2007). Higher temperatures should also lead to improved growth rates of the fish and together with increased recruitment is expected to lead to increased fish yields (Drinkwater, 2005; Stenevik and Sundby, 2007). The results of long-term simulations by the STOCOBAR model show that a temperature increase of $1-4 \mathrm{C}^{\circ}$ in the Barents Sea will lead not only to acceleration of cod growth and maturation rates, but also to increase in cannibalism (Stiansen et al. 2009). Increased overall production is expected to produce increased catches of cod, haddock and other species (ACIA, 2005). Cod are expected to spawn farther north and new spawning sites will likely be established (Sundby and Nakken, 2008; Drinkwater 2005). With increasing temperatures, temperate benthic species are expected to become more frequent and the species composition of the benthos will change. Such changes will affect benthic production (i.e. food for demersal fishes and other vertebrates) and may therefore have considerable management implications. Polar bears, ringed seals, bearded seals, harp seals and hooded seals are all dependent on sea ice. It is the primary foraging habitat for polar bears, and a resting and breeding habitat for all of these seals. Additionally, some of the seals feed on ice associated prey. As a result of climate warming and the associated loss of sea ice, distribution and abundance of these species are expected to decrease in the Barents Sea.
Along with climate change should mention that anthropogenic emissions of $\mathrm{CO}_{2}$ are causing acidification of the world oceans because $\mathrm{CO}_{2}$ reacts with seawater to form carbonic acid. Currently, acidity has increased by about $30 \%$ (reduction in pH by about 0,1 units). In 2100, pH reductions in the order of 0.2-0.3 units are predicted. This will significantly reduce the ability of organisms to build calcium carbonate shells and skeletons and it might also have other effects on organisms. The direct effects are expected to be most pronounced for phytoplankton, zooplankton and benthos. Fish, seabirds and marine mammals can be affected indirectly, possibly making ocean acidification one of the most important anthropogenic drivers in the Barents Sea in the future.

### 1.3 Description ofo the Barents Sea fisheries and its effect on the ecosystem (Tables 1.10-1.11, Figures 1.11-1.16

Description of the Barents Sea fisheries and its effect on the ecosystem (Tables 1.101.11, Figures 1.11-1.16)

Fishing is the largest human impact to the fish stocks in the Barents Sea, and thereby the functioning of the whole ecosystem. However, the observed variation in both fish species and ecosystem is also impacted by other effects such as climate and predation. Open ocean fisheries in the Barents Sea started in the beginning of the $20^{\text {th }}$ century with the development of trawling technology. At present there is a multinational fishery operating in the Barents Sea using different fishing gears and targeting several species. The largest commercially exploited fish stocks (Northeast Arctic cod, haddock and saithe) are now harvested within sustainable limits and have full reproduc-
tive capacity. However, some of the smaller stocks (golden redfish, beaked redfish and coastal cod) are overfished, and damage to benthic organisms and habitats from trawling has been documented. Overcoming these problems and further developing our understanding of the effects of fisheries in an ecosystem context are important challenges for management.

### 1.4 General description of the fisheries

The major demersal stocks in the Northeast Arctic include cod, haddock, saithe, and shrimp. In addition, redfish, Greenland halibut, wolffish, and flatfishes (e.g. long rough dab, plaice) are common on the shelf and at the continental slope, and ling and tusk at the slope and in deeper waters. In 2008, catches of nearly 900 thousand tonnes are reported from the stocks of cod, haddock, saithe, redfish, and Greenland halibut, which is an increase of $10 \%$ as compared to the year before. An additional catch of about 40000 tonnes was taken from the stocks of wolffish and shrimp. The annual fishing mortalities F (the mortality rate is linked to the proportion of the population being fished by $1-\mathrm{e}^{\mathrm{F}}$ ) for the assessed demersal fish stocks show large temporal variation within species and large differences across species from 0.1 ( $\approx 10 \%$ mortality) for some years for Sebastes marinus to above 1 ( $\approx 63 \%$ mortality) for some years for cod (Figure 1.11a). The current harvest rate relative to the maximum levels above which the fishing mortality over time may impair the recruitment is shown in Figure 1.11b. Of the analytically assessed demersal stocks in the Barents Sea it is currently only golden redfish (Sebastes marinus) which is harvested above this critical level.

The major pelagic stocks are capelin, herring, and polar cod. There was no fishery for capelin in the area in 2004-2008 due to the stock's poor condition, but in 2009 and 2010 the stock is again sufficiently sound to support a quota of 390000 and 360000 tonnes, respectively.
Russia, as the only nation currently fishing polar cod, fished 8190 tonnes polar cod in 2008. Norwegian spring spawning herring is the largest stock inhabiting the Northeast Arctic with its spawning stock estimated to 12.6 million tonnes in 2009.1 .5 million tonnes were fished from this stock in 2008, of which about 280000 tonnes were caught near the Norwegian coast in the south-western part of the Barents Sea. The highly migratory species blue whiting and mackerel extend their feeding migrations into this region, and in 2007 about 65000 tonnes mackerel and 120000 tonnes blue whiting were caught in the area, none of this, how ever, within the Barents Sea. Species with relatively small landings include salmon, Atlantic halibut, hake, pollack, whiting, Norway pout, anglerfish, lumpsucker, argentines, grenadiers, flatfishes, dogfishes, skates, crustaceans, and molluscs.

The most widespread gear used in the central Barents Sea is bottom trawl, but also long line and gillnets are used in the demersal fisheries. The pelagic fisheries use purse seine and pelagic trawl. Other gears more common along the coast include handline and Danish seine. Less frequently used gears are float line (used in a small but directed fishery for haddock along the coast of Finnmark, Norway) and various pots and traps for fish and crabs. The gears used vary with time, area and country, with Norway having the largest variety because of the coastal fishery. For Russia, the most common gear is bottom trawl, but a longline fishery mainly directed at cod and wolffish is also present. The other countries mainly use bottom trawl.

For most of the exploited stocks an agreed quota is decided (TAC), and also a number of additional regulations are applied. The regulations differ among gears and species and may be different from country to country, and a non-exhaustive list as well as a
description of the major fisheries in the Barents Sea by species can be found in Table 1.10.

From 2011 onwards, the minimum mesh size for bottom trawl fisheries for cod and haddock will be 130 mm for the entire Barents Sea (at present the minimum mesh size is 135 mm in the Norw egian EEZ and 125 mm in the Russian EEZ). This change is not expected to have a significant impact on the total exploitation pattern for these stocks, thus a recent average exploitation pattern is used in the predictions.

### 1.4.1 Mixed fisheries

The demersal fisheries are highly mixed, usually with a clear target species dominating, and with low linkage to the pelagic fisheries (Table 1.11). Although the degree of mixing may be high, the effect of the fisheries varies among the species. More specifically, the coastal cod stock and the two redfish stocks are presently at very low levels. Therefore, the effect of the mixed fishery will be largest for these stocks. In order to rebuild these stocks, further restrictions in the regulations should be considered (e.g. closures, moratorium, and restrictions in gears).

Successful management of an ecosystem includes being able to predict the effect of a mixed fishery on the individual stocks, and ICES is requested to provide advice which is consistent across stocks for mixed fisheries. Work on incorporating mixed fishery effects in ICES advice is ongoing and various approaches have been evaluated (ICES 2006/ACFM:14). At present such approaches are largely missing due to a need for improving methodology combined with lack of necessary data. How ever, technical interactions between the fisheries can be explored by the correlation in fishing mortalities among species (Figure 1.12). The correlation in fishing mortality is positive for Northeast Arctic cod and coastal cod, and for haddock and coastal cod confirming the linkage in these fisheries. There is also a significant relationship between saithe and Greenland halibut although the linkage in these fisheries is believed to be low (Table 1.11). The relationships between the other fishing mortalities are scattered and inconclusive. In case of strong dependencies in fishing mortalities this method can, in principle, be used to produce consistent advice across species concerning fishing mortality. It is however too simple since this correlation is influenced by too many confounding factors whose effect cannot be removed without a detailed analysis of data with a higher resolution (e.g. saithe and Greenland halibut, and changes in stock distribution (ICES 2006/ACFM:14).
A further quantification of the degree of mixing and impact on individual stocks requires detailed information about the target species and mix per catch/landing and gear. Such data exist for some fleets (e.g. the trawler fleet), but is incomplete for other fleets. The Russian and Norwegian trawl fleet catches show spatial and temporal differences in both composition and size as well as large differences between countries (Figures 1.13-1.16). In the north eastern part of the Barents Sea the major part of the Russian catches consists of cod, whereas the Norwegian catches include a large proportion of other species (mainly shrimp). In the most western part of the Barents Sea, the Norwegian catches consist of Sebastes mentella and Greenland halibut in addition to cod, whereas the Russian catches mainly consist of cod and haddock. The main reason for this disparity is the difference in spatial resolution of the data; the Norwegian strata system extends further west and thus covers the fishing grounds of Greenland halibut, whereas the Russian strata does not. The Norwegian trawl fishery along the Norw egian coast includes areas closer to the coast and is also more southerly distributed where other species are more dominant in the catches (e.g. saithe).

Estimates of unreported catches of cod and haddock in 2002-2008 indicate that this has been a considerable problem which now seems to be decreasing. According to the report from the Norw egian-Russian analytical group the total catches of both cod and haddock reported to AFWG are very close (within 1\%) to the estimates made by this group. Thus it was decided to set the IUU catches for 2009 to zero (see chapter 0.4). A continuous control and surveillance of this problem is necessary. Discarding of cod and haddock (and in some years also saithe) is thought to be significant in periods, although discarding of these, and a number of other species, is illegal in Norway and Russia. Data on discards are scarce, but attempts to obtain better quantification are ongoing.

## Fleet composition (groundfish and pelagic species)

Figure 1.17 shows the main fleets catching bottom and pelagic fishes in the Barents Sea and Svalbard (Spitsbergen archipelago) areas. The pelagic fishery is only conducted by Russia and Norway where both countries target the capelin. Russia has, in addition, fished polar cod with pelagic trawl (Norway has not fished this species since the early 1980s), and Norway has in recent years fished some legal sized herring in a restricted coastal purse seine fishery inside 4 nautical miles off Finnmark. Further in the south w estern part of the Barents Sea (south-w est of a line between Sørøya and Bear Island), extending into the Norwegian Sea, an international herring fishery has been open in some seasons.

The Norwegian groundfish fishery is much more diverse compared to Russia and other countries regarding the number of fleets. The trawler fleet itself is also rather diverse both within and between countries. In the Norwegian groundfish fishery several other gears are also used in addition to trawl. The gear composition also depends on which groundfish species the fishery targets. The Norwegian bottom trawl fleet catch about $30 \%$ of the Norwegian cod catch, about $40 \%$ of the haddock, and more than $40 \%$ of the Norwegian saithe and Greenland halibut catches. The Russian bottom trawl fleet catch about $100 \%$ of the Russian saithe catch, about $95 \%$ of cod and haddock, $90 \%$ of the Russian Greenland halibut catch and about $37 \%$ of wolffishes. Other countries fishing groundfish in these waters only use trawl, incl. some pairtrawling. It is mandatory in all groundfish trawl fisheries to use sorting grid to avoid catching undersized fish. The one and only exception from this rule is within an area in the southwestern part of the Barents Sea during 1 January - 30 April where trawling without sorting grids is permitted to catch haddock.

## Impact of fisheries on the ecosystem

In order to conclude on the total impact of trawling, an extensive mapping of fishing effort and bottom habitat would be necessary. In general, the response of benthic organisms to disturbance differs with substrate, depth, gear, and type of organism (Collie et al. 2000). Seabed characteristics from the Barents Sea are only scarcely known (Klages et al. 2004) and the lack of high-resolution ( $\pm 100 \mathrm{~m}$ ) maps of benthic habitats and biota is currently the most serious impediment to effective protection of vulnerable habitats from fishing activities (Hall 1999). An assessment of fishing intensity on fine spatial scales is critically important in evaluating the overall impact of fishing gear on different habitats and may be achieved, for example, by satellite tracking of fishing vessels (Jennings et al. 2000).The challenge for management is to determine levels of fishing that are sustainable and not degradable for benthic habitats in the longrun.

Fisheries in the Barents Sea do not only influence the targeted stocks. Due to strong species interactions fisheries removal of one stock may influence the abundance of other stocks. For example, herring collapses have positively influenced capelin abundance. Reduced stock sizes due to fisheries removal may also lead to changing migration patterns. Due to density dependent migrations, fish stocks cover greater areas and migrate longer distances when abundances are high compared to low. Fisheries also reduce the average fish size, age and age at maturity. The reduced size and age of the cod stock may actually have altered the ecological role of cod as top predators in the Barents Sea.

The qualitative effects of trawling have been studied to some degree. The most serious effects of otter trawling have been demonstrated for hard-bottom habitats dominated by large sessile fauna, where erected organisms such as sponges, anthozoans and corals have been shown to decrease considerably in abundance in the pass of the ground gear. Barents Sea hard bottom substrata, with associated attached large epifauna should therefore be identified.

Effects on soft bottom have been less studied, and consequently there are large uncertainties associated with what any effects of fisheries on thesehabitats might be. Studies on impacts of shrimp trawling on clay-silt bottoms have not demonstrated clear and consistent effects, but potential changes may be masked by the more pronounced temporal variability in these habitats (Løkkeborg 2005). The impacts of experimental trawling have been studied on a high seas fishing ground in the Barents Sea (Kutti et al. 2005.) Trawling seems to affect the benthic assemblage mainly through resuspension of surface sediment and through relocation of shallow burrowing infaunal species to the sur face of the seafloor.

Work is currently going on in the Arctic, jointly between Norway and Russia, exploring the possibility of using pelagic trawls when targeting demersal fish. The purpose is to avoid impact on bottom fauna and to reduce the mixture of other species. It will be mandatory to use sorting grids to avoid catches of undersized fish.

Lost gears such as gillnets may continue to fish for a long time (ghost fishing). The catch efficiency of lost gillnets has been examined for some species and areas (e.g. Humborstad et al. 2003; Misund et al. 2006; Large et al. 2009), but at present no estimate of the total effect is available. Ghost fishing in depths shallower than 200 m is usually not a significant problem because lost, discarded, and abandoned nets have a limited fishing life owing to their high rate of biofouling and, in some areas, their tangling by tidal scouring. Investigations made by the Norw egian Institute of Marine Research of Bergen in 1999 and 2000 showed that the amount of gillnets lost increases with depth and out of all the Norwegian gillnet fisheries, the Greenland halibut fishery is the metier where most nets are lost. The effect of ghost fishing in deeper water, e.g. for Greenland halibut, may be greater since such nets may continue to "fish" for periods of at least $2-3$ years, and perhaps even longer (D. M. Furevik and J. E. Fosseidengen, unpublished data), largely as a result of lesser rates of biofouling and tidal scouring in deep water. The Norwegian Directorate of Fisheries has organised retrieval surveys annually since 1980. All together 10784 gill nets of 30 metres standard length (approximately 320 km ) have been removed from Norwegian fishing grounds during the period from 1983 to 2003.

Other types of fishery-induced mortality include burst net, and mortality caused by contact with active fishing gear, such as escape mortality (Suuronen 2005; Broadhurst et al. 2006; Ingólfsson et al. 2007). Some small-scale effects are demonstrated, but the population effect is not known.

The harbour porpoise is common in the Barents Sea region south of the polar front and is most abundant in coastal waters. The harbour porpoise is subject to by-catches in gillnet fisheries (Bjerge and Kovacs 2005). In 2004 Norway initiated a monitoring program on by-catches of marine mammals in fisheries.
Fisheries impact seabird populations in two different ways: 1) Directly through bycatch of seabirds in fishing equipment and 2) Indirectly through competition with fisheries for the same food sources.

Documentation of the scale of by-catch of seabirds in the Barents Sea is fragmentary. Special incidents like the by-catch of large numbers of guillemots during spring cod fisheries in Norwegian areas have been documented (Strann et al. 1991). Gillnet fishing affects primarily coastal and pelagic diving seabirds, while the surface-feeding species will be most affected by long-line fishing (Furness 2003). The population impact of direct mortality through by-catch will vary with the time of year, the status of the affected population, and the sex and age structure of the birds killed. Even a numerically low by-catch may be a threat to red-listed species such as Common guillemot, Whitebilled diver and Steller's eider.

Several bird scaring devices has been tested for long-lining, and a simple one, the bird-scaring line (Løkkeborg 2003), not only reduces significantly bird by-catch, but also increases fish catch, as bait loss is reduced. This way there is an economic incentive for the fishermen to use it, and where bird by-catch is a problem, the bird-scaring line is used without any forced regulation.

In 2009, the Norwegian Institute for Nature Research (NINA) and the Institute of Marine Research in Norway started a cooperation to develop methods for estimation of bird by-catch. Preliminary reports from observers at sea trained by the institutes show that most of the fisheries have a minor impact on bird mortality.

### 1.5 Management improvement issues (Tables 1.12-1.15)

### 1.5.1 Overview

The availability of necessary ecosystem information is only one of the needed items for implementation of an ecosystem approach to management. Another needed element is the development of appropriate methods and instruments for incorporation of ecosystem information into stock assessment and harvest control rules.

This section summarizes ecosystem information that has the potential of being implemented in, and therefore improves, the advice for sustainable fishery management.

Management of fisheries is always based on decision-making under levels of uncertainty. Incorporating data on ocean climate, lower trophic level bio-production, as well as species interactions on higher trophic levels in catch recommendations for target species, should reduce the uncertainty of scientific recommendations for sustainable harvest levels.

### 1.5.2 Multispecies models

Development of multispecies models designed to improve fisheries management in the Barents Sea based on species interactions started in the mid-1980s. The first models developed were MULTSPEC, AGGMULT and SYSTMOD in IMR and MSVPA in PINRO (Tjelmeland and Bogstad, 1998; Hamre and Hatlebakk, 1998, Korzhev and Dolgov, 1999). In total, these models contained the species cod, capelin, herring, had-
dock, polar cod, shrimp, harp seal and minke whale. Even though further development of these models has been discontinued, they serve as predecessors to newly developed models, such as EcoCod, Bifrost, Gadget and STOCOBAR. Benefits of multispecies models include: improved estimates of natural mortality and recruitment; better understanding of stock-recruit relationships and variability in growth rates; alternatives views on biological reference points. Brief descriptions of the multispecies models are given below.

## EcoCod

The development of this model started in 2005 as the main task in the first stage of the joint PINRO-IMR Programme on Estimation of Maximum Long-Term Yield of North-East Arctic cod, taking into account the effect of ecosystem factors. This 10year research programme was initiated following a request from the RussianNorwegian Fishery Commission (Filin and Tjelmeland, 2005). EcoCod is a stepwise extension of a single species model for cod (CodSim; Kovalev and Bogstad, 2005), where cod growth, maturation, cannibalism and recruitment is modeled in a multispecies setting. Preliminary sub-models for cod growth, fecundity and malformation of eggs have been implemented in EcoCod.

## Bifrost

Bifrost (Boreal integrated fish resource optimization and simulation tool) is a multispecies model for the Barents Sea (Tjelmeland and Lindstrøm, 2005) with main emphasis on the cod-capelin dynamics. The prey items for cod are younger cod, capelin and other food. The predation model is estimated by comparing simulated consumption to that calculated from individual stomach content data using the dos Santos evacuation rate model with a parameterization where the initial meal size is excluded. The capelin availability partly shields the cod juveniles from cannibalism, and by including this effect, the recruitment relation for cod is significantly improved.

In prognostic mode, Bifrost is coupled to the assessment model for herring - SeaStar (Tjelmeland and Lindstrøm, 2005) - and the negative effect of herring juveniles on capelin recruitment is modeled through the recruitment function for capelin. Bifrost is also used to evaluate cod-capelin-herring multispecies harvest control rules.

## STOCOBAR

The STOCOBAR describes stock dynamics of cod in the Barents Sea, taking into account trophic interactions and environmental influence (Filin, 2007). It is designed as a tool for prediction and exploration of cod stock development as well as for evaluation of harvest strategies and recovery plans under different ecosystem scenarios. The STOCOBAR is an age-structured, a single-area and a single-fleet model with one year time steps. It includes a cod as predator on up to eight prey items: capelin, shrimp, polar cod, herring, krill, haddock, own young and other food. Species structure of the model is not permanent and it can be reduced from seven-species version to a simple version, which includes cod and capelin only. Recruitment function is used for cod only. Impact assessment of ecosystem factors on cod stock dynamics are based on «what if» scenarios. STOCOBAR is able to take uncertainties in future scenarios of temperature and capelin stock dynamics, in abundance and individual weight of cod at age 1 and in its fishing mortality rate into account. The first version of STOCOBAR was created at PINRO in 2001 and development of this model is continuing. The
work on the development of the STOCOBAR model is part of the Barents Sea Case Study within the EU project UNCOVER (2006-2010) and thejoint PINRO-IMR project (2004-2013) Optimal long-term harvest in the Barents Sea.

## GADGET

A multi-species Gadget age-length structured model (www hafro.is/gadget ; Begley and Howell, 2004, developed during the EU project dst² (2000-2003)), is being used for modeling the interactions between cod, herring, capelin and minke whale in the Barents Sea as part of the EU projects BECAUSE, UNCOVER, DEFINEIT and FACTS. This is a multi-area, multi-species model, focusing on predation interactions within the Barents Sea. The predator species are minke whale, cod and herring, with capelin, immature cod, and juvenile herring as prey species. Krill is included as an exogenous food for minke whales (Lindstrom et al. 2009). The cod model employed is based on the model presented at AFWG each year.

The modeling approach taken has many similarities to the MULTSPEC approach (Bogstad et al., 1997). Work is ongoing to enhance the modeling of recruitment processes during the EU projects FACTS and DEFINEIT. An FLR routine has been written that can run Gadget models as FLR Operating Models. This also gives the possibility of using Gadget as an operating model to test the performance of various assessment programs under a range of scenarios (Howell and Bogstad, 2010). In addition the Gadget multi-species model is being developed to assess the likely impact on medium-term population dynamics of oil-spill induced larval mortalities.

### 1.5.3 Statistical models

## Recruitment of commercial fish

Prediction of recruitment in fish stocks is essential for harvest prognosis stocks, both in a single-species and multi-species context. Traditionally, prediction methods have been based on spawning stock biomass and survey indices of juvenile fish and have not included effects of climate variability. Multiple linear regression models can be used to incor porate both climate and parental fish stock parameters. In order for such models to give predictions there need to be a a time lag between the predictor and response variables.

## Maturation of cod

The decrease in capelin stock biomass potentially impacts the maturation dynamics of Northeast Arctic cod by delaying the onset of maturation and/or increasing the incidence of skipped spawning. The relationship between weight- and length-at age shows that for a given length, weight-at-length is positively correlated with proportion mature-at-length for the period 1985-2001 (Marshall et al., 2004).

Estimates of weight-at-length were multiplied by the Russian liver condition index at length (Yaragina and Marshall, 2000) to derive estimates of liver weights in grams for cod at a standard length (see Marshall et al. 2004 for details of the calculation). This analysis indicated that for the period 1985-2001 there is a consistently significant, positive relationship between liver weight and proportion mature.

## Condition of fish

Relative body condition (the quantity of stored energy) is an important tool in understanding demographic variation and the ability of a population to respond to envi-
ronmental stressors, varying food availability and competition. A high-resolution database was used to examine causes of variation in the condition of North-east arctic cod for the period 1967-2004, over annual and monthly timescales. Temperature was shown to have a positive impact on condition at both inter- and intra-annual timescales. Interannually, temperature may affect stock distribution, in particular its overlap with the capelin stock. At shorter timescales it is likely that temperature directly affects the metabolism of the cod. Intra-annually, the quantity of capelin in cod stomachs positively affected cod condition in the current and the preceding month for all lengths of cod. This indicated a time lag between a change in food consumption and a subsequent change in condition, or 'latency'.

Results presented by Sandeman et al. (2008) point to the importance of the impact of varying temperature on condition. The effects of climate are likely to be particularly important where the species is close to its outer distribution area or where the animal is an ectotherm.

## Growth of fish

Large interannual variations in growth rate are observed for all commercial fish species in the Barents Sea. The most important causes are temperature change, density dependence and changes in prey availability. Variation in growth rate can contribute substantially to variability in stock biomass and can have a large impact on reproductive output. Regressions of weight at age of cod on temperature, capelin and the cod stock itself are used in EcoCod model.

Growth of the youngest capelin is correlated with abundance of the smallest zooplankton, whereas growth of older capelin is more closely correlated with abundance of the larger zooplankton. The developed regression equations have low determination coefficient, and are therefore not presented here. However, they may prove useful in the future when further developed.

## Reproductive potential

Morgan et al. (2009) explore the impact of four alternative indices of reproductive potential (RP) on perceptions of population productivity for eight fish populations across the North Atlantic. The four indices of RP included increasing biological complexity, adding variation in maturation, sex ratio, and fecundity. Perceptions of stock productivity were greatly affected by the choice of index of RP. Population status relative to reference points, RP per recruit, and projections of population size all varied when alternative indices of RP were used. There was no consistency in which index of RP gave the highest or lowest estimate of population productivity, but rather, this varied depending on how much variation there was in the reproductive biology of the population and the age composition. Estimates of sustainable harvest levels and recovery time for depleted populations can vary greatly depending on the index of RP.

### 1.5.4 Other models

## Consumption models

When calculating the prey consumption by a given predator, both the overall consumption level and the prey composition in the diet are used. The prey composition is usually derived from stomach content data, while the overall consumption level can be calculated using two a pproaches:

- A bioenergetic approach (as is usually the case for marine mammals and seabirds as predators)
- By combining data on stomach content weight with models for stomach evacuation rate, based on experiments.


## Ecosystem models

Ecosystem models may be useful for looking at how change in one species or ecosystem component is affecting whole or other parts an ecosystem, thereby identifying the most important inter-species/ functional group links and sensitivity of the ecosystem to changes to those. They are also useful for scenario testing (change in fishery pressure, climate change, and sudden pollution events. Special interesting are those models that have spatial resolution, like ATLANTIS and ECOPATH/ECOSIM.

Atlantis (Fulton et al., 2004a) is an ecosystem 3D box-model intended for use in management strategy evaluation (as described in de la Mare 1996, Cochrane et al. 1998, Butterw orth and Punt 1999, Sainsbury et al. 2000). The overall structure of Atlantis is based around having multiple alternative submodels to represent each step in the management strategy and adaptive management cycles. It has been applied to multiple marine systems (from single bays to millions of square kilometers) in Australia and the United States. In autumn 2010 it will be implemented at IMR, and cover the area of the Barents Sea and the Norwegian Sea.
Another model that may have some utility for the future would be ECOPATH/ECOSIM. This model can use ecosystem survey data and expected biomass conversion rates to model systems. As a mass-balance model it can detect if there may be overlooked components to the ecosystem. The ECOPATH model system is used in many systems around the world. Versions of it have also been applied to the Barents Sea ecosystem (Blanchard et al. 2002, Dommasnes et al. 2002), though they are not run on an operational level.

### 1.5.5 Expected impact of ecosystem factors on stock dynamics

## Evaluation of natural potential of cod stock biomass changes based on temperature and capelin data

STOCOBAR long-term simulations show that impact of capelin on cod stock dynamics is dependent on temperature and cod stock state (WD21). Using these simulations the natural potential for changes in cod stock size may be identified based on temperature conditions and the state of cod and capelin stocks in the Barents Sea. A table for evaluating the level of natural potential for annual changes in fishable cod stock biomass was produced based on the simulated data (Table1.12).

According to Table 1.12 and available data on temperature, cod and capelin stocks the potential for annual changes of cod fishable stock biomass in 2009 was low. The same situations will be in 2010 and 2011 based on expected temperature and capelin stock size. The resistance of cod stock to fishing pressure under these conditions will be medium and this does not imply high contributions to cod stock dynamics from capelin and temperature.

## Prediction of NEA cod recruitment.

Several statistical models, which use multiple linear regressions, have been developed for recruitment of North East Arctic cod. All models try to predict recruitment at age 3 (at 1 January), as calculated from the VPA, with cannibalism included. This quantity is denoted as R3. A collection of the most relevant models for AFWG is described below.

Stiansen et al. (2005) developed a model (JES1) with 2 year prediction possibility:

$$
\begin{aligned}
& \text { JES1: R3~ Temp(-3) + Age1(-2) + MatBio(-2) } \\
& \text { JES2: R3~ Temp(-3) + Age2(-1) + MatBio(-2) } \\
& \text { JES3: } \\
& \text { R3~ Temp(-3) + Age3(0) + MatBio(-2) }
\end{aligned}
$$

Temp is the Kola yearly temperature $(0-200 \mathrm{~m})$, Age1 is the winter survey bottom trawl index for cod age 1, and MatBio the maturing biomass of capelin. The number in parenthesis is the time lag in years. Two other similar models (JES2, JES3) can be made by substituting the term Age1 (-2) with Age2(-1) and Age3(0), respectively (winter survey bottom trawl index for cod age 2 and age 3, respectively), This gives 1 and 0 year predictions, respectively.

Svendsen et al. (2007) used a model (SV) based only data from the ROMS numerical hydro-dynamical model, with 3 year prognosis possibility:
SV: R3~ Phyto(-3) + Inflow (-3)

Where Phyto is the modelled phytoplankton production in the whole Barents Sea and Inflow is the modelled inflow through the western entrance to the Barents Sea in the autumn. The number in parenthesis is the time lag in years. The model have not been updated since 2007.

The recruitment model (TB) suggested by T. Bulgakova (AFWG 2005 WD14,WD9) is a modification of Ricker's model for stock-recruitment defined by:

$$
\text { TB: } \quad \text { R3 } \sim m(-3) \exp [-S S B(-3)+N(-3)]
$$

Where R3 is the number of age3 recruits for NEA cod, $m$ is an index of population fecundity, SSB is the spawning stock biomass and N is equal to the numbers of months with positive temperature anomalies (TA) on the Kola Section in the birth year for the year class. The number in parenthesis is the time lag in years. For the years before 1998 TA was calculated relatively to monthly average for the period 1951-2000. For intervals after 1998, the TA was calculated with relatively linear trend in the temperature for the period 1998-present. The model was run using two time intervals (using cod year classes 1984-2000 and year classes 1984-2004) for estimating the model coefficients. The models were not updated this year.

Titov (WD 22) and Titov et al. (WD 16 AFWG 2005) developed models with 1 to 4 year prediction possibility (TITOV1, TITOV2, TITOV3, TITOV4, respectively), based on the oxygen saturation at bottom layers of the Kola section stations 3-7 (OxSat), air temperature at the Murmansk station (Ta), water temperature: 3-7 stations of the Kola section (layer $0-200 \mathrm{~m}$ ) (Tw), ice coverage in the Barents Sea (I), spawning stock biomass (SSB), and the acoustic abundance of cod at age 1 and 2, derived from the joint winter Barents Sea acoustic survey:

TITOV 0: R3 ${ }^{1} \sim \mathrm{DOxSat}^{2}(\mathrm{t}-13)+\mathrm{DOxSat}(\mathrm{t}-13)+\mathrm{TTa}(\mathrm{t}-39)+\operatorname{CodA} 3(\mathrm{t}+1)+\mathrm{Tw}(\mathrm{t}-17)$
TITOV1: R3 ${ }^{1} \sim \mathrm{DOxSat}^{2}(\mathrm{t}-13)+\mathrm{DOxSat}(\mathrm{t}-13)+\mathrm{TTa}(\mathrm{t}-39)+\operatorname{CodA} 2(\mathrm{t}-11)+\mathrm{Tw}(\mathrm{t}-17)$

```
TITOV2: R3² \({ }^{2}\) DOxSat² \((\mathrm{t}-13)-\mathrm{DOxSat}(\mathrm{t}-13)+\mathrm{ITa}(\mathrm{t}-39)+\operatorname{CodA} 1(\mathrm{t}-23)+\mathrm{Tw}(\mathrm{t}-17)\)
TITOV3: \(\mathrm{R}^{3}{ }^{3} \sim \mathrm{OxSat}^{2}(\mathrm{t}-44)+\mathrm{Ta}(\mathrm{t}-39)+\lg \operatorname{CodC0}(\mathrm{t}-28)\)
TITOV4: R3 \({ }^{4} \sim\) OxSat \({ }^{2}(\mathrm{t}-44)+\mathrm{Ta}(\mathrm{t}-39)+\mathrm{SSB}(\mathrm{t}-36)\)
```

Where DOxSat(t-13) $\operatorname{Exp}($ OxSat(t-13)) - OxSat(t-38), ITa(t-39) ~I(t-39) +Ta(t -44).The number in parenthesis is the time lag in months, relative to 1 January at age 3 . The ITa index coincides in time with the increase of horizontal gradients of water temperatures in the area of the Polar Front (Titov, 2001). Some changes were brought in 2009 (AFWG 2009 WD 12). New equation (TITOV0) was added, 0-group abundance indices, corrected for capture efficiency (CodC0) was entered instead of former indices in TITOV3.

Hjermann et al. (2007) developed a model with a one year prognosis, which have been modified by Dingsør et al (WD19) to four models with 1-2 year projection possibility.

$$
\begin{aligned}
& \text { H1: } \log (\mathrm{R} 3) \sim \operatorname{Temp}(-3)+\log (\text { Age } 0)(-3)+\mathrm{BM}_{\text {cod3-6 }} / \mathrm{ABM}_{\text {capelin }}(-2,-1) \\
& \text { H2: } \log (\mathrm{R} 3) \sim \operatorname{Temp}(-2)+\mathrm{I}(\text { surv })+\operatorname{Age} 1(-2)+\mathrm{BM}_{\text {cod3-6 }} / \mathrm{ABM}_{\text {capelin }}(-2,-1) \\
& \text { H3: } \log (\mathrm{R} 3) \sim \operatorname{Temp}(-1)+\text { Age2(-1) }+\mathrm{BM}_{\text {cod3-6 }} / \mathrm{ABM}_{\text {capelin }}(-1) \\
& \text { H4: } \log (\mathrm{R} 3) \sim \operatorname{Temp}(-1)+\text { Age3(0) }
\end{aligned}
$$

Temp is the Kola yearly temperature $(0-200 \mathrm{~m})$, Age0 is the 0 -group index of cod, Age1, Age2 and Age3 are the winter survey bottom trawl index for cod age 1,2 and 3, respectively, $\mathrm{BM}_{\text {cod } 36}$ is the biomass of cod between age 3 and 6 , and ABM is the maturing biomass of capelin. The number in parenthesis is the time lag in years.
At AFWG 2008, Subbey et al. presented a comparative study (AFWG 2008 WD27) on the ability of some of the above models in predicting stock recruitment for NEA cod (Age 3). At the assessment meeting this year a WD by Dingsør et al. (WD 19) was presented, which investigated the performance of some of the mentioned recruitment models. Even though this work was well received by the working group it was decided not to change the procedure this year. However, it was strongly recommended that a Study Group should be appointed to look at criteria's for choosing/rejecting recruitment models suitable for use in stock assessment (see also chap 0.11).

The 2008 assessment agreed on using a combination of the best performing models according to Subbey at (AFWG 2008 WD27) for the age 3 predictions, names the "hybrid" model. One-year-ahead prognoses was given by the hybrids (Titov1, Titov3 and JES1), two-year-ahead (Titov2, Titov 3 and JES1) and three-year-ahead (Titov3) for the number of age 3 cod. Following the recommendation of the review group in 2008 this procedure was also conducted in the 2009 assessment.

At the 2010 assessment the model JES 1 was removed from the hybrid for the 2010 estimate only, due to a low age 1 index and thereby the model being out of its valid range for that prognosis year. Otherwise the hybrid model approach was similar to last year.

Table 1.13 show the estimates of all the available models, along with last year estimates.

## Cannibalism mortality for cod

Currently AFWG estimates of cod natural mortality caused by cannibalism based on data of the cod proportion in the cod diet is shown in Table 1.14. These data are used for estimation of cod consumed by cod and further for estimation of its natural mor-
tality within the XSA (see section 3.4.2). Averaged natural mortality for last 3 years is used as predicted $M$ for next 4 years (section 3.7.1).

An alternative approach for prediction of NEA cod cannibalism was proposed by Kovalev (2004), based on the linear relationship between the natural mortality of cod at ages 3-5 and the biomass of cod spawning stock with minus 3-year lag. Using this approach the predicted natural mortality coefficient for cod, including cannibalism for recent years, seems to be higher compared to "the standard" assessment and prediction (Table 1.15). Because the mechanisms of cod SSB influence on the level of own young natural mortality on age 3-4 years is unclear, and because of this relationship seems not to be in correspondence with observations over the last few years, the assessment group decided that this approach should not to be used for prediction before it will be further tested. Values for the years 2009 to 2012, predicted by the regression, are given in the Table1.15.

### 1.5.6 Fishery induced evolution

There is a vital need for the fisheries science community to maintain sustainable fisheries ensuring the effective conservation, management and development of living aquatic resources. The precautionary approach was proclaimed and applied within the ICES community to meet (promote) these aims. This approach takes into account uncertainties relating to the size and productivity of the stocks. Uncertainties relating to fisheries induced evolution are most likely taken into consideration in case of a proper implementation of precautionary approach into responsible fishery.

The Study Group on Fisheries Induced Adaptive Change (SGFIAC) proposed to create evolutionary impact assessment (EvoIA), quantifying the evolutionary effects of management measures (ICES 2008/RMC:01; ICES 2009/RMC:03). It is a very complicated but promising task given that commercial fishery could act as a selective factor resulting in evolutionary response of exploited populations.

The papers published by the SGFIAC Group members concern basically probabilistic maturation reaction norms (PMRNs) estimations for different commercial stocks/species, and shift in cohort-specific PMRNs interpreted as a genetic change at the population level. It is rather difficult to test that findings directly as the genes associated with maturation have a polygenic nature. The strength and weakness of the PMRNs approach were discussed in detail in Theme Session issue of the Marine ecology progress series, 2007, vol. 335, 249- 310.

North east arctic cod stock demonstrates long-term trends in maturation as well as in demography of the stock and weight at length of fish. The historical trends could be caused both by genetic and plastic effects on maturation. Population density factors and environmental conditions can contribute to feeding success resulting in changing maturation rates in NEA cod for the time period investigated (Marshall and McAdam, 2007; Kovalev and Yaragina, 2009). The causes in a discontinuity of the decreasing trend observed in length for $50 \%$ maturation probability in the beginning of the 80's are unknown, but they are most likely non-genetic given that they occurred synchronously across age-classes (Marshall and McAdam, 2007).

More research is needed to evaluate underlined mechanisms of population changes including biological, physiological, ecological studies, not to mention genetic ones.

It takes a lot of time and efforts for the ICES community to implement the precautionary approach into a scientific/management practice. It is likely to take some time before the SGFIAC can evaluate and present some results applicable to test on real
management measures recommendations. AFWG considers it premature at present to discuss any proposals of management measures (or reference points for fisheries management) in terms of fisheries induced evolution. Dialogues with scientists of the mentioned WG could also be carried out through the ICES Sharepoint.

### 1.6 Monitoring of the ecosystem (Figure 1.18, Tables 1.16-1.17)

Monitoring of the Barents Sea started already in 1900 (initiated by Nicolai Knipovich), with regular measurement of temperature in the Kola section. In the last 50 years regular observations of ecosystem components in the Barents Sea have been conducted both at sections and by area covering surveys from ship and airplanes. In addition, there are conducted many long and short time special investigations, designed to study specific processes or knowledge gaps. Also, the quality of large hydrodynamical numeric models is now at a level where they are useful for filling observation gaps in time and space for some parameters. Satellite data and hindcast global reanalysed datasets are also useful information sources.

### 1.6.1 Standard sections and fixed stations

Some of the longest ocean time series in the w orld are along standard sections (Figure 1.18) in the Barents Sea. The monitoring of basic oceanographic variables for most of the sections goes back 30-50 years, with the longest time series stretching over one century. In the last decades also zooplankton is sampled at some of these sections. An overview of length, observation frequency and present measured variables for the standard sections in the Barents Sea is given in Table 1.16.

IMR operates one fixed station, Ingøy, related to the Barents Sea. The Ingøy station is situated in the coastal current along the Norwegian coast. Temperature and salinity is monitored 1-4 times a month. The observations were obtained in two periods, 19361944 and 1968-present.

### 1.6.2 Area coverage

Area surveys are conducted throughout the year. The number of vessels in each survey differs, not only between surveys but may also change from year to year for the same survey. However, most surveys are conducted with only one vessel. It is not possible to measure all ecosystem components during each survey. Effort is always put on measuring as many parameters as possible on each survey, but available time put restrictions on what is possible to accomplish. Also, an investigation should not take too long time in order to give a synoptic picture of the conditions. Therefore the surveys must focus on a specific set of parameters/species. Other measured parameters may therefore not have optimal coverage and thereby increased uncertainty, but will still give important information. An overview of the measured parameters/species on each main survey is given in Table 1.16. Specific considerations for the most important surveys are given in the following text.

## Norwegian / Russian winter survey

The survey is carried out during February-early March, and covers the main cod distribution area in the Barents Sea. The coverage is in some years limited by the ice distribution. Three vessels are normally applied, two Norwegian and one Russian. The main observations are made with bottom trawl, pelagic trawl, echo sounder and CTD. Plankton studies have been done in some years. Cod and haddock are the main targets for this survey. Swept area indices are calculated for cod, haddock, Greenland
halibut, S. marinus and S. mentella. Acoustic observations are made for cod, haddock, capelin, redfish, polar cod and herring. The survey started in 1981.

## Lofoten survey

The main spawning grounds of North East Arctic cod are in the Lofoten area. Echosounder equipment was first used in 1935 to detect concentrations of spawning cod, and the first attempt to map such concentrations was made in 1938 (Sund, 1938). Later investigations have provided valuable information on the migratory patterns, the geographical distribution and the age composition and abundance of the stock.
The current time series of survey data starts in 1985. Due to the change in echo sounder equipment in 1990 results obtained earlier are not directly comparable with later results. The survey is designed as equidistant parallel acoustic transects covering 3 strata (North, South and Vestfjorden). In most surveys previous to 1990 the transects were not parallel, but more as parts of a zig-zag pattern across the spawning grounds aimed at mapping the distribution of cod. Trawl samples are not taken according to a proper trawl survey design. This is due to practical reasons. The spawning concentrations can be located with echosounder thus effectively reduce the number of trawl stations needed. The ability to properly sample the composition of the stock (age, sex, maturity stage etc.) is limited by the amount of fixed gear (gillnets and longlines) in the different areas.

## Norwegian coastal surveys

In 1985-2002 a Norwegian acoustic survey specially designed for saithe was conducted annually in October-November (Nedreaas 1998). The survey covered the near coastal banks from the Varangerfjord close to the Russian border and southwards to $62^{\circ} \mathrm{N}$. The whole area has been covered since 1992, and the major parts since 1988. The aim of conducting an acoustic survey targeting Northeast Arctic saithe was to support the stock assessment with fishery-independent data of the abundance of the youngest saithe. The survey mainly covered the grounds where the trawl fishery takes place, normally dominated by $3-5(6)$ year old fish. 2-year-old saithe, mainly inhabiting the fiords and more coastal areas, were also represented in the survey, although highly variable from year to year. In 1995-2002 a Norwegian acoustic survey for coastal cod was conducted along the coast and in the fjords from Varanger to Stad in September, just prior to the saithe survey described above. This survey covered coastal areas not included in the regular saithe survey. Autumn 2003 the saithe- and coastal cod surveys were combined. The survey now also covers 0 -group herring in fjords north of Lofoten.

## Joint ecosystem autumn survey

The survey is carried out from early August to early October, and covers the whole Barents Sea. Four or five vessels are normally applied, three Norwegian and one or two Russian. Most aspects of the ecosystem are covered, from physical and chemical oceanography, primary and secondary production, fish (both young and adult stages), sea mammals, benthos and birds. Many kinds of methods and gears are used, water sampling, plankton nets, pelagic and demersal trawls, grabs and sledges, acoustics, directs observations (birds and sea mammals). The survey has developed from joint surveys on 0-group, capelin and juvenile Greenland halibut, through general acoustic surveys including observations of physical oceanography and plankton, gradually developing into the ecosystem survey carried out in recent years. The pre-
decessor of the survey dates back to 1972 and has been carried out every fall since. From 2003 these surveys were called "ecosystem surveys".

In 2009 not all components of the ecosystem were covered during the survey, and a further reduction will probably take place in 2010; the coverage of e.g. Greenland halibut will be less complete than in previous years. Also, the future of this ecosystem survey is still undetermined.
Associated with this survey Russia also covers parts of the Northern Kara Sea during autumn.

## Russian Autumn-winter trawl-acoustic survey

The survey is carried out in October-December, and cover the whole Barents Sea up to the continental slope. Two Russian vessels are usually used. The survey has developed from a young cod and haddock trawl survey, started in 1946. The current trawlacoustic time series of survey data starts in 1984, targeting both young and adult stages of bottom fish. The surveys include observations of physical oceanography and meso- and macro-zooplankton.

## Norwegian Greenland halibut survey

The survey is carried out in August, and cover the continental slope from 68 to $80^{\circ} \mathrm{N}$, in depths of $400-1500 \mathrm{~m}$ north of $70^{\circ} 30^{\prime} \mathrm{N}$, and $400-1000 \mathrm{~m}$ south of this latitude. This survey was run the first time in 1994, and is now part of the Norwegian Combined survey index for Greenland halibut. This survey will not be conducted in 2010, and its future design is being revised.

## Russian young herring survey

This survey is conducted in May and takes 2-3 weeks. It is including also observations of physical oceanography and plankton. In 1991-1995 it was joint survey, since 1996 the survey is carried out only by PINRO.

### 1.6.3 Other information sources

Large 3D hydrodynamic numeric models for the Barents Sea are run at both IMR and PINRO. These models have, through validation with observations, proved to be a useful tool for filling observation gaps in time and space. The hydrodynamic models have also proved useful for scenario testing, and for study of drift patterns of various planktonic organisms.

Sub-models for phytoplankton and zooplankton are now implemented in some of the hydrodynamic models. However, due to the present assumptions in these submodels care must be taken in the interpretation of the model results.
Satellites can be for several monitoring tasks. Ocean color specter can be used to identify and estimate the amount of phytoplankton in the skin ( $\sim 1 \mathrm{~m}$ ) layer. Several climate variables can be monitored (e.g. ice cover, cloud cover, heat radiation, sea surface temperature). Marine mammals, polar bears and seabirds can be traced with attached transmitters.

Aircraft surveys also are used for monitoring several physical parameters associated with the sea surface as well as observations of mammals at the surface and estimations of harp seal pup production in the White Sea.

Several international hindcast databases (e.g.. NCEP, ERA40) are available. They use a combination of numerical models and available observations to estimate several climate variables, covering the whole world.

Along the Norwegian coast ship-of-opportunity supply weekly the surface temperature along their path.

### 1.7 Main conclusions

State and expected situation in the ecosystem (section 1.2)

## Climate

- The air temperature was above the long-term mean during 2009.
- The sea temperature in the Barents Sea is still high, and about the same level as in 2008. There was an increase in the end of the year, with the highest December temperature in the Kola section. In 2010 the temperature is expected to further decrease, but still be higher than the long-term mean.
- Salinity in 2009 is still high, and at about the same levels as in 2008
- Inflow of Atlantic waters at the western entrance in 2009 was quite similar to 2008; moderate during winter followed by a strong decrease in spring. Data for second half of 2009 is not available.
- Oxygen levels were about normal in 2009.
- Ice extent in 2008 was less than normal, but more than in 2008. In 2010 ice conditions is expected to be slightly less or around the long term mean.


## Plankton and northern shrimp

- The mesozooplankton biomass measured in August-September 2009 was less compared to 2008, and below the long-term mean.
- Abundance euphausiids (krill) in autumn and winter 2009 were close to the level in 2008 in the western and northwestern areas and increased in the centre, eastern and coastal areas. In total the abundance in 2010 is slightly above the long-term mean.
- The abundance of gelatinous zooplankton, caught by pelagic trawling, show a lower abundance in 2009 compared to 2008.
- The shrimp stock in the Barents Sea and Spitsbergen area in 2009 increased compared to both 2007 and 2008.


## Fish

- Capelin stock size is at around average level, with a slight decrease from last year. The survey estimate at age 1 of the 2008 year class is slightly below the long-term mean. 0-group estimates indicate that the 2009 year class is below average.
- For young herring there are indications that the year classes 2005-2009 are below average. Therefore the abundance of herring in the Barents Sea is believed to be at a relatively low level in 2010.
- Blue whiting is still at a very low level, with a slight increase from 2008. The abundance is expected to remain low in 2009.
- The polar cod stock is presently at a high level, similar to 2008.


## Harp Seal

- The decrease in the harp seal pup production in the White Sea has become slower recently and even some slight increase has been observed, but it is still at a low level.


## Impact of fisheries on the ecosystem (section 1.3)

- The most widespread gear is trawl.
- The demersal fisheries are mixed, and currently have largest effect on coastal cod and redfish due to the poor condition of these stocks.
- The pelagic fisheries are less mixed, and are weakly linked to the demersal fisheries (however, by-catches of young pelagic stages of demersal species have been reported in some pelagic fisheries)
- Trawling has largest effect on hard bottom habitats; whereas the effects on other habitats are not clear and consistent.
- Work is currently going on exploring the possibility of using pelagic trawls when targeting demersal fish. The purpose is to avoid impact on bottom fauna and to reduce the mixture of other species. It will be mandatory to use sorting grids to avoid catches of undersized fish.
- Fishery induced mortality (lost gillnets, contact with active fishing gears, etc.) on fish is a potential problem but not quantified at present.
Management improvement iss ues (section 1.4)
- Several methods, which take ecosystem information into account, are presently under development. These methods should in the future be valuable for the improvement of the stock assessment and advice.
- According to STOCOBAR simulations there is a low probability to expect any tendency of decline or increase in the fishable cod stock biomass in 2010 and 2011, based on predicted temperature and capelin stock size.
- The cod recruitment (age 3) in 2010 is expected to be low compared to the long-term mean. In 2011 and 2012 it is expected to increase slightly, but still be below the long-term mean.


### 1.8 Response to technical minutes

There were no specific comments from the review group to ecosystem consideration chapter (Chapter 1 ).

Table 1.1. 0 -group abundance indices (in millions) with $95 \%$ confidence limits, not corrected for catching efficiency

| Year | Capelin |  |  | Cod |  |  | Haddock |  |  | Herring |  |  | Redfish |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  |
| 1980 | 197278 | 131674 | 262883 | 72 | 38 | 105 | 59 | 38 | 81 | 4 | 1 | 8 | 277873 | 0 | 701273 |
| 1981 | 123870 | 71852 | 175888 | 48 | 33 | 64 | 15 | 7 | 22 | 3 | 0 | 8 | 153279 | 0 | 363283 |
| 1982 | 168128 | 35275 | 300982 | 651 | 466 | 835 | 649 | 486 | 812 | 202 | 0 | 506 | 106140 | 63753 | 148528 |
| 1983 | 100042 | 56325 | 143759 | 3924 | 1749 | 6099 | 1356 | 904 | 1809 | 40557 | 19526 | 61589 | 172392 | 33352 | 311432 |
| 1984 | 68051 | 43308 | 92794 | 5284 | 2889 | 7679 | 1295 | 937 | 1653 | 6313 | 1930 | 10697 | 83182 | 36137 | 130227 |
| 1985 | 21267 | 1638 | 40896 | 15484 | 7603 | 23365 | 695 | 397 | 992 | 7237 | 646 | 13827 | 412777 | 40510 | 785044 |
| 1986 | 11409 | 98 | 22721 | 2054 | 1509 | 2599 | 592 | 367 | 817 | 7 | 0 | 15 | 91621 | 0 | 184194 |
| 1987 | 1209 | 435 | 1983 | 167 | 86 | 249 | 126 | 76 | 176 | 2 | 0 | 5 | 23747 | 12740 | 34755 |
| 1988 | 19624 | 3821 | 35427 | 507 | 296 | 718 | 387 | 157 | 618 | 8686 | 3325 | 14048 | 107027 | 23378 | 190675 |
| 1989 | 251485 | 201110 | 301861 | 717 | 404 | 1030 | 173 | 117 | 228 | 4196 | 1396 | 6996 | 16092 | 7589 | 24595 |
| 1990 | 36475 | 24372 | 48578 | 6612 | 3573 | 9651 | 1148 | 847 | 1450 | 9508 | 0 | 23943 | 94790 | 52658 | 136922 |
| 1991 | 57390 | 24772 | 90007 | 10874 | 7860 | 13888 | 3857 | 2907 | 4807 | 81175 | 43230 | 119121 | 41499 | 0 | 83751 |
| 1992 | 970 | 105 | 1835 | 44583 | 24730 | 64437 | 1617 | 1150 | 2083 | 37183 | 21675 | 52690 | 13782 | 0 | 36494 |
| 1993 | 330 | 125 | 534 | 38015 | 15944 | 60086 | 1502 | 911 | 2092 | 61508 | 2885 | 120131 | 5458 | 0 | 13543 |
| 1994 | 5386 | 0 | 10915 | 21677 | 11980 | 31375 | 1695 | 825 | 2566 | 14884 | 0 | 31270 | 52258 | 0 | 121547 |
| 1995 | 862 | 0 | 1812 | 74930 | 38459 | 111401 | 472 | 269 | 675 | 1308 | 434 | 2182 | 11816 | 3386 | 20246 |
| 1996 | 44268 | 22447 | 66089 | 66047 | 42607 | 89488 | 1049 | 782 | 1316 | 57169 | 28040 | 86299 | 28 | 8 | 47 |
| 1997 | 54802 | 22682 | 86922 | 67061 | 49487 | 84634 | 600 | 420 | 780 | 45808 | 21160 | 70455 | 132 | 0 | 272 |
| 1998 | 33841 | 21406 | 46277 | 7050 | 4209 | 9890 | 5964 | 3800 | 8128 | 79492 | 44207 | 114778 | 755 | 23 | 1487 |
| 1999 | 85306 | 45266 | 125346 | 1289 | 135 | 2442 | 1137 | 368 | 1906 | 15931 | 1632 | 30229 | 46 | 14 | 79 |
| 2000 | 39813 | 1069 | 78556 | 26177 | 14287 | 38068 | 2907 | 1851 | 3962 | 49614 | 3246 | 95982 | 7530 | 0 | 16826 |
| 2001 | 33646 | 0 | 85901 | 908 | 152 | 1663 | 1706 | 1113 | 2299 | 844 | 177 | 1511 | 6 | 1 | 10 |
| 2002 | 19426 | 10648 | 28205 | 19157 | 11015 | 27300 | 1843 | 1276 | 2410 | 23354 | 12144 | 34564 | 130 | 20 | 241 |
| 2003 | 94902 | 41128 | 148676 | 17304 | 10225 | 24383 | 7910 | 3757 | 12063 | 28579 | 15504 | 41653 | 216 | 0 | 495 |
| 2004 | 16701 | 2541 | 30862 | 19157 | 13987 | 24328 | 19144 | 12649 | 25638 | 133350 | 94873 | 171826 | 849 | 0 | 1766 |
| 2005 | 41808 | 12316 | 71300 | 21532 | 14732 | 28331 | 33283 | 24377 | 42190 | 26332 | 1132 | 51532 | 12332 | 631 | 24034 |
| 2006 | 166400 | 102749 | 230050 | 7860 | 3658 | 12061 | 11421 | 7553 | 15289 | 66819 | 22759 | 110880 | 20864 | 10057 | 31671 |
| 2007 | 157913 | 87370 | 228456 | 9707 | 5887 | 13527 | 2826 | 1787 | 3866 | 22481 | 4556 | 40405 | 159159 | 44882 | 273436 |
| 2008 | 288799 | 178860 | 398738 | 52975 | 31839 | 74111 | 2742 | 830 | 4655 | 15915 | 4477 | 27353 | 9962 | 0 | 20828 |
| 2009 | 189767 | 113154 | 266379 | 54579 | 37311 | 71846 | 13040 | 7988 | 18093 | 18916 | 8249 | 29582 | 66671 | 29636 | 103706 |
| Mean | 77706 |  |  | 19880 |  |  | 4040 |  |  | 28579 |  |  | 64744 |  |  |

Table 1.1. (cont.). 0 -group abundance indices (in millions) with $95 \%$ confidence limits, not corrected for catching efficiency.

| Year | Saithe |  | Gr halibut |  | Long rough dab |  | Polar cod (east) |  | Polar cod (west) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Abundance index | Confidence limit | Abundance index | Confidence limit | Abundance index | Confidence limit | Abundance index | Confidence limit | Abundance index | Confidence limit |
| 1980 | 3 | $0 \quad 6$ | 111 | $35 \quad 187$ | 1273 | 8831664 | 28958 | 978448132 | 9650 | $0 \quad 20622$ |
| 1981 | 0 | 00 | 74 | $46 \quad 101$ | 556 | $300 \quad 813$ | 595 | $226 \quad 963$ | 5150 | 19568345 |
| 1982 | 143 | 0371 | 39 | $11 \quad 68$ | 1013 | 6981328 | 1435 | 1442725 | 1187 | 03298 |
| 1983 | 239 | $83 \quad 394$ | 41 | $22 \quad 59$ | 420 | 264577 | 1246 | 02501 | 9693 | $0 \quad 20851$ |
| 1984 | 1339 | $407 \quad 2271$ | 31 | $18 \quad 45$ | 60 | $43 \quad 77$ | 127 | 0303 | 3182 | 7375628 |
| 1985 | 12 | $1 \quad 23$ | 48 | $29 \quad 67$ | 265 | $110 \quad 420$ | 19220 | 498933451 | 809 | $0 \quad 1628$ |
| 1986 | 1 | $0 \quad 2$ | 112 | $60 \quad 164$ | 6846 | 49418752 | 12938 | $2355 \quad 23521$ | 2130 | 1804081 |
| 1987 | 1 | $0 \quad 1$ | 35 | $23 \quad 47$ | 804 | $411 \quad 1197$ | 7694 | 017552 | 74 | $31 \quad 117$ |
| 1988 | 17 | 430 | 8 | 313 | 205 | 113297 | 383 | $9 \quad 757$ | 4634 | $0 \quad 9889$ |
| 1989 | 1 | $0 \quad 3$ | 1 | 03 | 180 | $100 \quad 260$ | 199 | $0 \quad 423$ | 18056 | 218233931 |
| 1990 | 11 | 20 | 1 | 02 | 55 | $26 \quad 84$ | 399 | $129 \quad 669$ | 31939 | $0 \quad 70847$ |
| 1991 | 4 | 26 | 1 | $0 \quad 2$ | 90 | $49 \quad 131$ | 88292 | 39856136727 | 38709 | $0 \quad 110568$ |
| 1992 | 159 | $86 \quad 233$ | 9 | $0 \quad 17$ | 121 | $25 \quad 218$ | 7539 | $0 \quad 15873$ | 9978 | 159118365 |
| 1993 | 366 | 0913 | 4 | 27 | 56 | $25 \quad 87$ | 41207 | 096068 | 8254 | 135915148 |
| 1994 | 2 | 0 5 | 39 | 093 | 1696 | 10832309 | 267997 | 151917384078 | 5455 | $0 \quad 12032$ |
| 1995 | 148 | $68 \quad 229$ | 15 | $5 \quad 24$ | 229 | $39 \quad 419$ | 1 | $0 \quad 2$ | 25 | 149 |
| 1996 | 131 | $57 \quad 204$ | 6 | $3 \quad 9$ | 41 | 279 | 70134 | 4319697072 | 4902 | $0 \quad 12235$ |
| 1997 | 78 | $37 \quad 120$ | 5 | $3 \quad 7$ | 97 | $44 \quad 150$ | 33580 | 1878848371 | 7593 | 62314563 |
| 1998 | 86 | $39 \quad 133$ | 8 | 312 | 27 | $13 \quad 42$ | 11223 | $6849 \quad 15597$ | 10311 | 023358 |
| 1999 | 136 | $68 \quad 204$ | 14 | $8 \quad 21$ | 105 | 1210 | 129980 | 82936177023 | 2848 | 4075288 |
| 2000 | 206 | 111301 | 43 | $17 \quad 69$ | 233 | $120 \quad 346$ | 116121 | 67589164652 | 22740 | 1492430556 |
| 2001 | 20 | $0 \quad 46$ | 51 | $20 \quad 83$ | 162 | $78 \quad 246$ | 3697 | $658 \quad 6736$ | 13490 | $0 \quad 28796$ |
| 2002 | 553 | 108998 | 51 | $0 \quad 112$ | 731 | $342 \quad 1121$ | 96954 | 57530136378 | 27753 | 418451322 |
| 2003 | 65 | $0 \quad 146$ | 13 | $0 \quad 34$ | 78 | $45 \quad 110$ | 11211 | 610016323 | 1627 | $0 \quad 3643$ |
| 2004 | 1395 | $860 \quad 1930$ | 70 | $28 \quad 113$ | 36 | $20 \quad 52$ | 37156 | 1904055271 | 367 | $125 \quad 610$ |
| 2005 | 55 | $36 \quad 73$ | 9 | $4 \quad 14$ | 200 | 109292 | 6540 | $3196 \quad 9884$ | 3216 | 12695162 |
| 2006 | 142 | $60 \quad 224$ | 11 | 120 | 710 | $437 \quad 983$ | 26016 | 999642036 | 2078 | $464 \quad 3693$ |
| 2007 | 51 | $6 \quad 96$ | 1 | 0 | 262 | $45 \quad 478$ | 25883 | 849443273 | 2532 | $0 \quad 5134$ |
| 2008 | 45 | $22 \quad 69$ | 6 | $0 \quad 13$ | 956 | $410 \quad 1502$ | 6649 | 84512453 | 91 | $0 \quad 183$ |
| 2009 | 22 | $0 \quad 46$ | 7 | 410 | 115 | $51 \quad 179$ | 23570 | 966137479 | 21433 | 564237223 |
| Mean | 181 |  | 29 |  | 587 |  | 35898 |  | 8997 |  |

Table 1.2. 0 -group abundance indices (in millions) with $95 \%$ confidence limits, corrected for catching efficiency.

| Year | Capelin |  |  | Cod |  |  | Haddock |  |  | Herring |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  |
| 1980 | 740289 | 49518 | 985391 | 276 | 131 | 421 | 265 | 169 | 361 | 77 | 12 | 142 |
| 1981 | 477260 | 27349 | 681026 | 289 | 201 | 377 | 75 | 34 | 117 | 37 | 0 | 86 |
| 1982 | 599596 | 14529 | 1053893 | 3480 | 2540 | 4421 | 2927 | 2200 | 3655 | 2519 | 0 | 5992 |
| 1983 | 340200 | 19112 | 489278 | 19299 | 9538 | 29061 | 6217 | 3978 | 8456 | 195446 | 69415 | 321477 |
| 1984 | 275233 | 16140 | 389057 | 24326 | 14489 | 34164 | 5512 | 3981 | 7043 | 27354 | 3425 | 51284 |
| 1985 | 63771 | 589 | 121648 | 66630 | 32914 | 100346 | 2457 | 1520 | 3393 | 20081 | 3933 | 36228 |
| 1986 | 41814 | 64 | 82986 | 10509 | 7719 | 13299 | 2579 | 1621 | 3537 | 93 | 27 | 160 |
| 1987 | 4032 | 145 | 6607 | 1035 | 504 | 1565 | 708 | 432 | 984 | 49 | 0 | 111 |
| 1988 | 65127 | 1210 | 118153 | 2570 | 1519 | 3622 | 1661 | 630 | 2693 | 60782 | 20877 | 100687 |
| 1989 | 862394 | 69098 | 1033806 | 2775 | 1624 | 3925 | 650 | 448 | 852 | 17956 | 8252 | 27661 |
| 1990 | 115636 | 7730 | 153966 | 23593 | 13426 | 33759 | 3122 | 2318 | 3926 | 15172 | 0 | 36389 |
| 1991 | 169455 | 7407 | 264832 | 40631 | 29843 | 51419 | 13713 | 10530 | 16897 | 267644 | 107990 | 427299 |
| 1992 | 2337 | 25 | 4423 | 166276 | 92113 | 240438 | 4739 | 3217 | 6262 | 83909 | 48399 | 119419 |
| 1993 | 952 | 28 | 1616 | 133046 | 58312 | 207779 | 3785 | 2335 | 5236 | 291468 | 1429 | 581506 |
| 1994 | 13898 |  | 27725 | 70761 | 39933 | 101589 | 4470 | 2354 | 6586 | 103891 | 0 | 212765 |
| 1995 | 2869 |  | 6032 | 233885 | 114258 | 353512 | 1203 | 686 | 1720 | 11018 | 4409 | 17627 |
| 1996 | 136674 | 6980 | 203546 | 280916 | 188630 | 373203 | 2632 | 1999 | 3265 | 549608 | 256160 | 843055 |
| 1997 | 189372 | 8073 | 298011 | 294607 | 218967 | 370247 | 1983 | 1391 | 2575 | 463243 | 176669 | 749817 |
| 1998 | 113390 | 7051 | 156263 | 24951 | 15827 | 34076 | 14116 | 9524 | 18707 | 476065 | 277542 | 674589 |
| 1999 | 287760 | 14324 | 432278 | 4150 | 944 | 7355 | 2740 | 1018 | 4463 | 35932 | 13017 | 58848 |
| 2000 | 140837 | 655 | 275123 | 108093 | 58416 | 157770 | 10906 | 6837 | 14975 | 469626 | 22507 | 916746 |
| 2001 | 90181 |  | 217345 | 4150 | 798 | 7502 | 4649 | 3189 | 6109 | 10008 | 2021 | 17996 |
| 2002 | 67130 | 3697 | 97288 | 76146 | 42253 | 110040 | 4381 | 2998 | 5764 | 151514 | 58954 | 244073 |
| 2003 | 340877 | 14617 | 535575 | 81977 | 47715 | 116240 | 30792 | 15352 | 46232 | 177676 | 52699 | 302653 |
| 2004 | 53950 | 1199 | 95900 | 65969 | 47743 | 84195 | 39303 | 26359 | 52246 | 773891 | 544964 | 1002819 |
| 2005 | 148466 | 5166 | 245263 | 72137 | 50662 | 93611 | 91606 | 67869 | 115343 | 125927 | 20407 | 231447 |
| 2006 | 515770 | 32577 | 705764 | 25061 | 11469 | 38653 | 28505 | 18754 | 38256 | 294649 | 102788 | 486511 |
| 2007 | 480069 | 27231 | 687825 | 42628 | 26652 | 58605 | 8401 | 5587 | 11214 | 144002 | 25099 | 262905 |
| 2008 | 995101 | 62720 | 1362999 | 234144 | 131081 | 337208 | 9864 | 1144 | 18585 | 201046 | 68778 | 333313 |
| 2009 | 673027 | 42338 | 922668 | 185457 | 123375 | 247540 | 33339 | 19707 | 46970 | 104233 | 31009 | 177458 |
| Mean | 266916 |  |  | 76659 |  |  | 11243 |  |  | 169164 |  |  |

Table 1.2 (cont.). 0-group abundance indices (in millions) with $95 \%$ confidence limits, corrected for catching efficiency.

| Year | Saithe |  | Polar cod (east) |  |  | Polar cod (west) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Abundance index | Confidence limit | Abundance index | Confide | nce limit | Abundance index | Confiden | ce limit |
| 1980 | 21 | 47 | 203226 | 69898 | 336554 | 82871 | 0 | 176632 |
| 1981 | 0 | $0 \quad 0$ | 4882 | 1842 | 7922 | 46155 | 17810 | 74500 |
| 1982 | 296 | 0699 | 1443 | 154 | 2731 | 10565 | 0 | 29314 |
| 1983 | 562 | 211912 | 1246 | 0 | 2501 | 87272 | 0 | 190005 |
| 1984 | 2577 | 7254430 | 871 | 0 | 2118 | 26316 | 6097 | 46534 |
| 1985 | 30 | $7 \quad 53$ | 143257 | 39633 | 246881 | 6670 | 0 | 13613 |
| 1986 | 4 | $0 \quad 9$ | 102869 | 16336 | 189403 | 18644 | 125 | 37164 |
| 1987 | 4 | $0 \quad 10$ | 64171 | 0 | 144389 | 631 | 265 | 996 |
| 1988 | 32 | $11 \quad 52$ | 2588 | 59 | 5117 | 41133 | 0 | 89068 |
| 1989 | 10 | $0 \quad 23$ | 1391 | 0 | 2934 | 164058 | 15439 | 312678 |
| 1990 | 29 | 455 | 2862 | 879 | 4846 | 246819 | 0 | 545410 |
| 1991 | 9 | 414 | 823828 | 366924 | 1280732 | 281434 | 0 | 799822 |
| 1992 | 326 | 156495 | 49757 | 0 | 104634 | 80747 | 12984 | 148509 |
| 1993 | 1033 | 02512 | 297397 | 0 | 690030 | 70019 | 12321 | 127716 |
| 1994 | 7 | 12 | 2139223 | 1230225 | 3048220 | 49237 | 0 | 109432 |
| 1995 | 415 | 196634 | 6 | 0 | 14 | 195 | 0 | 390 |
| 1996 | 430 | $180 \quad 679$ | 588020 | 368361 | 807678 | 46671 | 0 | 116324 |
| 1997 | 341 | 162521 | 297828 | 164107 | 431550 | 62084 | 6037 | 118131 |
| 1998 | 182 | $91 \quad 272$ | 96874 | 59118 | 134630 | 95609 | 0 | 220926 |
| 1999 | 275 | 139411 | 1154149 | 728616 | 1579682 | 24015 | 3768 | 44262 |
| 2000 | 851 | 4461256 | 916625 | 530966 | 1302284 | 190661 | 133249 | 248072 |
| 2001 | 47 | $0 \quad 106$ | 29087 | 5648 | 52526 | 119023 | 0 | 252146 |
| 2002 | 2112 | 1344090 | 829216 | 496352 | 1162079 | 215572 | 36403 | 394741 |
| 2003 | 286 | 0631 | 82315 | 42707 | 121923 | 12998 | 0 | 30565 |
| 2004 | 4779 | $2810 \quad 6749$ | 290686 | 147492 | 433879 | 2892 | 989 | 4796 |
| 2005 | 176 | $115 \quad 237$ | 44663 | 22890 | 66436 | 25970 | 9987 | 41953 |
| 2006 | 280 | 116443 | 182713 | 73645 | 291781 | 15965 | 3414 | 28517 |
| 2007 | 286 | 3568 | 191111 | 57403 | 324819 | 22803 | 0 | 46521 |
| 2008 | 142 | $68 \quad 216$ | 42657 | 5936 | 79378 | 619 | 25 | 1212 |
| 2009 | 62 | $0 \quad 132$ | 168990 | 70509 | 267471 | 154687 | 37022 | 272351 |
| Mean | 520 |  | 291798 |  |  | 73411 |  |  |

Table 1.3. The North-east arctic cod stock's consumption of various prey species in 1984-2009 (1000 tonnes), based on Norwegian consumption calculations.

| Year | Other | Amphipods | Krill | Shrimp | Capelin | Herring | Polar cod | Cod | Haddock | Redfish | G. halibut | $\begin{array}{r} \text { Blue } \\ \text { whiting } \end{array}$ | Long rough dab | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 479 | 27 | 113 | 436 | 722 | 78 | 15 | 22 | 50 | 364 | 0 | 0 | 24 | 2330 |
| 1985 | 1112 | 170 | 58 | 156 | 1621 | 183 | 3 | 32 | 47 | 225 | 0 | 1 | 41 | 3649 |
| 1986 | 606 | 1236 | 111 | 142 | 837 | 133 | 141 | 83 | 111 | 315 | 0 | 0 | 55 | 3769 |
| 1987 | 671 | 1085 | 67 | 191 | 229 | 32 | 206 | 25 | 4 | 324 | 1 | 0 | 9 | 2844 |
| 1988 | 401 | 1237 | 318 | 129 | 339 | 8 | 92 | 9 | 3 | 223 | 0 | 4 | 5 | 2769 |
| 1989 | 656 | 800 | 241 | 131 | 572 | 3 | 32 | 8 | 10 | 228 | 0 | 0 | 57 | 2739 |
| 1990 | 1343 | 137 | 85 | 195 | 1609 | 7 | 6 | 19 | 15 | 243 | 0 | 87 | 95 | 3842 |
| 1991 | 760 | 65 | 76 | 188 | 2891 | 8 | 12 | 26 | 20 | 312 | 7 | 10 | 270 | 4646 |
| 1992 | 907 | 102 | 158 | 373 | 2457 | 331 | 97 | 55 | 106 | 188 | 20 | 2 | 93 | 4889 |
| 1993 | 751 | 253 | 714 | 315 | 3033 | 163 | 278 | 285 | 71 | 100 | 2 | 2 | 26 | 5994 |
| 1994 | 625 | 563 | 704 | 518 | 1085 | 147 | 582 | 224 | 49 | 79 | 0 | 1 | 39 | 4614 |
| 1995 | 845 | 982 | 516 | 362 | 628 | 115 | 254 | 371 | 116 | 193 | 1 | 0 | 34 | 4417 |
| 1996 | 599 | 631 | 1158 | 341 | 538 | 47 | 104 | 536 | 69 | 97 | 0 | 10 | 34 | 4164 |
| 1997 | 443 | 382 | 519 | 316 | 907 | 5 | 113 | 338 | 41 | 36 | 0 | 33 | 14 | 3146 |
| 1998 | 411 | 363 | 455 | 325 | 714 | 86 | 151 | 155 | 33 | 9 | 0 | 13 | 15 | 2730 |
| 1999 | 378 | 145 | 271 | 250 | 1720 | 128 | 220 | 62 | 26 | 16 | 1 | 31 | 7 | 3255 |
| 2000 | 385 | 167 | 464 | 450 | 1727 | 53 | 194 | 76 | 51 | 8 | 0 | 38 | 18 | 3633 |
| 2001 | 685 | 172 | 376 | 277 | 1722 | 71 | 250 | 66 | 49 | 6 | 1 | 151 | 29 | 3853 |
| 2002 | 362 | 96 | 261 | 232 | 1934 | 86 | 270 | 108 | 123 | 1 | 0 | 224 | 15 | 3713 |
| 2003 | 548 | 282 | 529 | 240 | 2157 | 214 | 272 | 114 | 168 | 3 | 0 | 74 | 48 | 4649 |
| 2004 | 671 | 679 | 318 | 247 | 1296 | 196 | 338 | 122 | 193 | 3 | 12 | 74 | 62 | 4212 |
| 2005 | 685 | 411 | 521 | 264 | 1238 | 187 | 354 | 116 | 342 | 2 | 3 | 111 | 46 | 4282 |
| 2006 | 780 | 169 | 957 | 313 | 1511 | 201 | 118 | 70 | 361 | 15 | 1 | 122 | 104 | 4721 |
| 2007 | 1141 | 293 | 935 | 373 | 1881 | 272 | 228 | 94 | 355 | 40 | 1 | 39 | 61 | 5712 |
| 2008 | 1384 | 146 | 787 | 316 | 2443 | 102 | 476 | 182 | 303 | 55 | 11 | 29 | 90 | 6325 |
| 2009 | 1250 | 159 | 427 | 211 | 2762 | 219 | 478 | 175 | 295 | 35 | 1 | 5 | 91 | 6109 |

Table 1.4. The North-east arctic COD stock's consumption of various prey species in 1984-2009 (1000 tonnes), based on Russian consumption calculations.

| Year |  | $\begin{aligned} & \text { 总 } \\ & \text { 2 } \\ & \text { 会 } \end{aligned}$ | $\begin{gathered} \frac{0}{\pi} \\ \frac{\pi}{\omega} \\ \hline \end{gathered}$ | $\begin{aligned} & 0.0 \\ & \text { E } \\ & \text { E } \\ & \text { T } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { E } \\ & \text { ט. } \\ & \text { Ũ } \\ & \hline \end{aligned}$ |  | $\stackrel{\rightharpoonup}{0}$ |  |  | $$ | $\begin{aligned} & \frac{\pi}{0} \\ & \stackrel{y}{0} \\ & \underset{\sim}{2} \end{aligned}$ |  |  | $\begin{aligned} & \frac{\pi}{n} \\ & \vdots \\ & \pm \\ & 0 \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \\ & \pm \\ & 0 \\ & \hline \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 93 | 31 | 351 | 33 | 592 | 17 | 13 | 50 | 5 | 1 | 195 | 51 | 0 | 269 | 286 | 1988 |
| 1985 | 30 | 432 | 202 | 24 | 989 | 0 | 98 | 34 | 18 | 15 | 97 | 23 | 0 | 519 | 198 | 2679 |
| 1986 | 57 | 860 | 148 | 47 | 807 | 159 | 28 | 103 | 3 | 27 | 158 | 24 | 1 | 372 | 170 | 2962 |
| 1987 | 69 | 508 | 201 | 8 | 162 | 105 | 27 | 2 | 10 | 15 | 118 | 6 | 0 | 268 | 188 | 1686 |
| 1988 | 209 | 169 | 118 | 19 | 292 | 0 | 20 | 93 | 0 | 0 | 127 | 20 | 0 | 239 | 242 | 1545 |
| 1989 | 167 | 290 | 104 | 4 | 680 | 34 | 34 | 2 | 0 | 0 | 158 | 56 | 0 | 201 | 248 | 1977 |
| 1990 | 101 | 30 | 270 | 64 | 1254 | 8 | 21 | 16 | 39 | 15 | 232 | 79 | 0 | 101 | 167 | 2396 |
| 1991 | 54 | 83 | 287 | 28 | 3286 | 44 | 52 | 22 | 7 | 6 | 144 | 46 | 6 | 132 | 158 | 4354 |
| 1992 | 213 | 38 | 263 | 374 | 2021 | 191 | 84 | 38 | 0 | 77 | 121 | 44 | 1 | 295 | 418 | 4175 |
| 1993 | 186 | 177 | 223 | 177 | 2791 | 171 | 147 | 153 | 4 | 25 | 41 | 48 | 5 | 160 | 384 | 4691 |
| 1994 | 362 | 298 | 472 | 105 | 1303 | 492 | 391 | 72 | 1 | 2 | 56 | 40 | 0 | 100 | 353 | 4046 |
| 1995 | 396 | 465 | 550 | 192 | 691 | 203 | 557 | 130 | 0 | 1 | 113 | 53 | 3 | 169 | 356 | 3878 |
| 1996 | 973 | 361 | 200 | 76 | 478 | 79 | 473 | 60 | 9 | 37 | 71 | 47 | 0 | 470 | 175 | 3509 |
| 1997 | 386 | 85 | 207 | 54 | 523 | 110 | 409 | 35 | 3 | 0 | 37 | 33 | 2 | 97 | 399 | 2380 |
| 1998 | 615 | 205 | 265 | 70 | 852 | 129 | 129 | 23 | 23 | 18 | 15 | 19 | 0 | 53 | 226 | 2641 |
| 1999 | 454 | 77 | 242 | 74 | 1402 | 165 | 48 | 14 | 25 | 1 | 13 | 8 | 0 | 58 | 107 | 2688 |
| 2000 | 413 | 111 | 367 | 48 | 1662 | 157 | 57 | 29 | 26 | 8 | 4 | 20 | 0 | 36 | 181 | 3119 |
| 2001 | 418 | 74 | 308 | 88 | 1433 | 140 | 59 | 49 | 137 | 29 | 4 | 31 | 2 | 145 | 190 | 3106 |
| 2002 | 309 | 45 | 198 | 55 | 2330 | 281 | 100 | 77 | 102 | 3 | 4 | 17 | 0 | 44 | 170 | 3734 |
| 2003 | 240 | 140 | 213 | 144 | 1155 | 204 | 127 | 323 | 26 | 5 | 1 | 38 | 0 | 87 | 270 | 2974 |
| 2004 | 350 | 378 | 243 | 122 | 1046 | 350 | 83 | 151 | 48 | 20 | 7 | 58 | 15 | 179 | 267 | 3317 |
| 2005 | 543 | 135 | 226 | 170 | 962 | 318 | 114 | 275 | 68 | 42 | 7 | 45 | 2 | 162 | 203 | 3272 |
| 2006 | 887 | 62 | 210 | 239 | 1186 | 108 | 95 | 268 | 104 | 86 | 17 | 95 | 1 | 92 | 333 | 3781 |
| 2007 | 860 | 153 | 280 | 259 | 1408 | 239 | 73 | 319 | 33 | 21 | 22 | 65 | 1 | 194 | 376 | 4304 |
| 2008 | 617 | 36 | 229 | 102 | 2324 | 498 | 138 | 333 | 16 | 16 | 42 | 109 | 13 | 301 | 416 | 5191 |
| 2009 | 511 | 105 | 199 | 158 | 2380 | 575 | 115 | 306 | 7 | 82 | 27 | 185 | 0 | 133 | 510 | 5293 |
| Mean | 366 | 206 | 253 | 105 | 1308 | 184 | 134 | 114 | 27 | 21 | 70 | 48 | 2 | 188 | 269 | 3296 |

Table 1.5. Consumption per cod by cod age group ( $\mathrm{kg} / \mathrm{year}$ ), based on Norwegian consumption calculations.

| Year/Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0.247 | 0.814 | 1.684 | 2.513 | 3.948 | 5.203 | 7.973 | 8.486 | 9.139 | 9.867 | 9.941 |
| 1985 | 0.304 | 0.761 | 1.829 | 3.101 | 4.671 | 7.357 | 11.172 | 11.892 | 12.416 | 13.660 | 13.773 |
| 1986 | 0.160 | 0.488 | 1.347 | 3.158 | 5.604 | 6.834 | 10.989 | 11.899 | 12.701 | 13.461 | 13.694 |
| 1987 | 0.219 | 0.601 | 1.275 | 2.055 | 3.537 | 5.457 | 7.044 | 8.111 | 8.922 | 9.343 | 9.295 |
| 1988 | 0.164 | 0.703 | 1.149 | 2.148 | 3.744 | 5.875 | 10.096 | 11.218 | 12.570 | 13.122 | 13.345 |
| 1989 | 0.223 | 0.716 | 1.606 | 2.705 | 3.973 | 5.601 | 7.648 | 8.464 | 9.559 | 10.156 | 10.599 |
| 1990 | 0.363 | 0.905 | 1.889 | 3.027 | 4.156 | 5.323 | 6.249 | 6.666 | 6.698 | 7.039 | 7.675 |
| 1991 | 0.293 | 0.969 | 2.168 | 3.500 | 5.281 | 7.026 | 9.392 | 10.154 | 11.200 | 12.239 | 11.886 |
| 1992 | 0.215 | 0.663 | 2.095 | 3.133 | 4.142 | 5.093 | 7.832 | 8.965 | 9.352 | 10.071 | 10.115 |
| 1993 | 0.112 | 0.528 | 1.546 | 3.044 | 4.809 | 6.285 | 9.421 | 11.239 | 11.763 | 12.253 | 12.876 |
| 1994 | 0.130 | 0.408 | 0.922 | 2.521 | 3.504 | 4.511 | 6.396 | 8.846 | 9.672 | 9.977 | 10.176 |
| 1995 | 0.103 | 0.296 | 0.921 | 1.840 | 3.361 | 5.252 | 7.697 | 10.405 | 12.333 | 12.734 | 13.180 |
| 1996 | 0.108 | 0.356 | 0.929 | 1.847 | 3.068 | 4.429 | 7.381 | 11.143 | 14.702 | 14.876 | 15.265 |
| 1997 | 0.140 | 0.319 | 0.940 | 1.768 | 2.710 | 3.536 | 5.253 | 8.149 | 12.582 | 13.484 | 13.091 |
| 1998 | 0.117 | 0.397 | 0.983 | 1.942 | 2.923 | 4.186 | 5.746 | 8.061 | 11.339 | 11.850 | 11.903 |
| 1999 | 0.163 | 0.505 | 1.093 | 2.717 | 3.717 | 5.442 | 6.965 | 9.179 | 11.004 | 12.007 | 12.109 |
| 2000 | 0.170 | 0.499 | 1.243 | 2.461 | 4.252 | 5.651 | 7.951 | 9.364 | 12.485 | 13.258 | 13.299 |
| 2001 | 0.171 | 0.456 | 1.309 | 2.439 | 3.682 | 5.294 | 7.523 | 11.085 | 13.422 | 14.117 | 14.434 |
| 2002 | 0.199 | 0.551 | 1.167 | 2.441 | 3.380 | 4.719 | 6.357 | 9.039 | 10.224 | 11.538 | 10.921 |
| 2003 | 0.207 | 0.653 | 1.312 | 2.390 | 3.995 | 5.946 | 8.411 | 10.405 | 12.786 | 13.397 | 14.343 |
| 2004 | 0.194 | 0.474 | 1.280 | 2.529 | 3.882 | 5.588 | 7.323 | 11.213 | 16.665 | 18.557 | 17.980 |
| 2005 | 0.194 | 0.653 | 1.376 | 2.592 | 3.918 | 5.588 | 7.182 | 9.771 | 13.090 | 14.012 | 14.784 |
| 2006 | 0.181 | 0.595 | 1.589 | 2.796 | 4.185 | 5.870 | 7.482 | 11.255 | 13.695 | 14.692 | 15.613 |
| 2007 | 0.213 | 0.621 | 1.742 | 3.178 | 4.704 | 6.231 | 7.802 | 9.621 | 12.636 | 13.223 | 13.808 |
| 2008 | 0.189 | 0.665 | 1.460 | 3.047 | 4.336 | 6.667 | 8.135 | 10.842 | 14.166 | 14.673 | 14.883 |
| 2009 | 0.182 | 0.586 | 1.414 | 2.888 | 4.568 | 5.789 | 8.074 | 10.195 | 12.252 | 13.203 | 13.286 |
| Average | 0.191 | 0.584 | 1.396 | 2.606 | 4.001 | 5.564 | 7.822 | 9.835 | 11.821 | 12.589 | 12.780 |

Table 1.6. Consumption per cod by cod age group ( $\mathrm{kg} / \mathrm{year}$ ), based on Russian consumption calculations.

| Year/Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0.262 | 0.893 | 1.612 | 2.748 | 3.848 | 5.486 | 6.990 | 8.563 | 10.574 | 13.166 | 12.437 | 14.282 | 15.272 |
| 1985 | 0.295 | 0.752 | 1.656 | 2.683 | 4.264 | 6.601 | 8.242 | 9.743 | 10.975 | 14.447 | 16.499 | 16.061 | 17.343 |
| 1986 | 0.179 | 0.515 | 1.461 | 3.467 | 4.956 | 5.913 | 6.477 | 8.156 | 9.766 | 11.455 | 12.500 | 13.577 | 14.772 |
| 1987 | 0.145 | 0.431 | 0.844 | 1.561 | 3.078 | 4.346 | 7.279 | 9.683 | 12.703 | 14.482 | 15.014 | 15.115 | 16.377 |
| 1988 | 0.183 | 0.704 | 1.075 | 1.627 | 2.392 | 4.387 | 8.208 | 9.978 | 10.867 | 16.536 | 14.352 | 15.765 | 12.361 |
| 1989 | 0.282 | 0.910 | 1.468 | 2.207 | 3.244 | 4.799 | 6.581 | 8.725 | 11.134 | 15.799 | 15.950 | 17.909 | 14.023 |
| 1990 | 0.288 | 1.007 | 1.696 | 2.694 | 3.278 | 3.833 | 5.584 | 6.871 | 10.716 | 11.428 | 12.660 | 15.053 | 16.064 |
| 1991 | 0.241 | 0.936 | 2.670 | 4.473 | 6.038 | 7.846 | 9.590 | 11.542 | 14.970 | 19.294 | 17.509 | 20.109 | 22.109 |
| 1992 | 0.178 | 0.969 | 2.475 | 2.866 | 3.995 | 5.138 | 6.724 | 7.414 | 8.754 | 12.304 | 13.518 | 13.744 | 14.908 |
| 1993 | 0.133 | 0.476 | 1.512 | 2.865 | 3.944 | 5.108 | 7.372 | 8.945 | 10.343 | 11.600 | 14.067 | 14.893 | 15.922 |
| 1994 | 0.180 | 0.512 | 1.212 | 2.402 | 3.517 | 5.359 | 7.560 | 10.001 | 11.818 | 12.896 | 13.554 | 15.902 | 16.806 |
| 1995 | 0.194 | 0.497 | 0.962 | 1.819 | 3.204 | 4.847 | 7.332 | 9.688 | 13.835 | 15.247 | 16.960 | 18.230 | 19.202 |
| 1996 | 0.170 | 0.498 | 1.028 | 1.916 | 3.075 | 4.189 | 6.987 | 10.212 | 12.185 | 13.426 | 14.581 | 16.214 | 16.876 |
| 1997 | 0.119 | 0.341 | 0.992 | 1.908 | 2.668 | 3.503 | 4.954 | 7.980 | 12.174 | 21.523 | 20.666 | 21.822 | 24.237 |
| 1998 | 0.232 | 0.528 | 1.081 | 2.016 | 2.823 | 4.089 | 5.469 | 7.346 | 9.586 | 13.012 | 14.455 | 15.579 | 16.201 |
| 1999 | 0.261 | 0.431 | 1.128 | 2.490 | 3.676 | 5.222 | 6.398 | 8.220 | 9.194 | 13.364 | 15.325 | 16.918 | 17.567 |
| 2000 | 0.186 | 0.545 | 1.288 | 2.551 | 4.387 | 6.559 | 8.833 | 10.483 | 11.522 | 15.132 | 17.155 | 19.717 | 20.514 |
| 2001 | 0.150 | 0.413 | 1.163 | 2.110 | 3.430 | 5.571 | 6.835 | 10.233 | 12.457 | 15.130 | 17.374 | 19.322 | 20.559 |
| 2002 | 0.252 | 0.677 | 1.303 | 2.699 | 3.847 | 5.591 | 7.846 | 10.796 | 13.238 | 18.787 | 17.902 | 20.202 | 21.207 |
| 2003 | 0.228 | 0.618 | 1.296 | 2.028 | 3.547 | 4.716 | 6.684 | 8.905 | 13.418 | 14.492 | 19.540 | 19.239 | 20.036 |
| 2004 | 0.250 | 0.654 | 1.412 | 2.567 | 3.857 | 5.660 | 7.730 | 11.126 | 15.907 | 20.770 | 21.687 | 24.852 | 25.892 |
| 2005 | 0.255 | 0.687 | 1.514 | 2.504 | 3.896 | 5.264 | 7.192 | 9.395 | 13.163 | 15.981 | 22.656 | 23.387 | 24.181 |
| 2006 | 0.354 | 0.921 | 1.833 | 2.763 | 3.986 | 5.317 | 7.396 | 10.202 | 12.762 | 16.462 | 21.563 | 25.940 | 26.875 |
| 2007 | 0.234 | 0.666 | 1.803 | 3.018 | 4.295 | 5.810 | 7.444 | 9.017 | 11.754 | 15.961 | 20.903 | 25.154 | 26.064 |
| 2008 | 0.223 | 0.706 | 1.641 | 2.881 | 4.071 | 6.006 | 7.705 | 10.317 | 13.471 | 17.596 | 22.968 | 27.431 | 27.328 |
| 2009 | 0.217 | 0.627 | 1.503 | 2.542 | 4.266 | 5.530 | 7.617 | 10.986 | 13.258 | 15.637 | 21.532 | 25.632 | 25.586 |

Table 1.7. Capelin stock history from 1973-present. M output biomass is the estimated biomass of capelin removed from the stock by natural mortality.

| Year | Total stock number, billions (Oct. 1) | Total stock biomass in 1000 tonnes (Oct. 1) | Maturing biomass in 1000 tonnes (Oct. 1) | M output biomass (MOB) during year (1000 tonnes) |
| :---: | :---: | :---: | :---: | :---: |
| 1973 | 961 | 5144 | 1350 | 5504 |
| 1974 | 1029 | 5733 | 907 | 4542 |
| 1975 | 921 | 7806 | 2916 | 4669 |
| 1976 | 696 | 6417 | 3200 | 5633 |
| 1977 | 681 | 4796 | 2676 | 4174 |
| 1978 | 561 | 4247 | 1402 | 3782 |
| 1979 | 464 | 4162 | 1227 | 5723 |
| 1980 | 654 | 6715 | 3913 | 5708 |
| 1981 | 660 | 3895 | 1551 | 5658 |
| 1982 | 735 | 3779 | 1591 | 3729 |
| 1983 | 754 | 4230 | 1329 | 3884 |
| 1984 | 393 | 2964 | 1208 | 3051 |
| 1985 | 109 | 860 | 285 | 1975 |
| 1986 | 14 | 120 | 65 | 681 |
| 1987 | 39 | 101 | 17 | 200 |
| 1988 | 50 | 428 | 200 | 80 |
| 1989 | 209 | 864 | 175 | 537 |
| 1990 | 894 | 5831 | 2617 | 415 |
| 1991 | 1016 | 7287 | 2248 | 3307 |
| 1992 | 678 | 5150 | 2228 | 7745 |
| 1993 | 75 | 796 | 330 | 4631 |
| 1994 | 28 | 200 | 94 | 982 |
| 1995 | 17 | 193 | 118 | 163 |
| 1996 | 96 | 503 | 248 | 261 |
| 1997 | 140 | 911 | 312 | 828 |
| 1998 | 263 | 2056 | 931 | 915 |
| 1999 | 285 | 2776 | 1718 | 2070 |
| 2000 | 595 | 4273 | 2099 | 2464 |
| 2001 | 364 | 3630 | 2019 | 3906 |
| 2002 | 201 | 2210 | 1290 | 2939 |
| 2003 | 104 | 533 | 280 | 3195 |
| 2004 | 82 | 628 | 293 | 812 |
| 2005 | 42 | 324 | 174 | 817 |
| 2006 | 88 | 787 | 437 | 733 |
| 2007 | 280 | 1885 | 836 | 2033 |
| 2008 | 570 | 4426 | 2468 | 3285 |
| 2009 | 352 | 3756 | 2322 | * |

[^0]Table 1.8. Diet composition of main fish species in 2005, \% by weight (Data from Dolgov, WD 28 and WD 29, AFWG 2006)

| PREY SPECIES | Predators species |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \hline \text { Cod } \\ (3+) \end{gathered}$ | haddock | Greenland halibut | Thorny skate | Long rough dab | Saithe | $\begin{array}{r} \text { Blue } \\ \text { whiting } \end{array}$ |
| Euphausiidae | 5,2 | 21,7 | 0,4 | 0,8 | 0,1 | 24,4 | 44,4 |
| Hy periidae | 4,1 | 0,2 | 3,8 | 0 | 0 | 0,3 | 18,2 |
| Cephalopoda | 0 | 0 | 2,1 | 0 | 0 | 0 | 0 |
| Pandalus borealis | 4,6 | 1,2 | 1,4 | 15,8 | 1,4 | 0,2 | 1,4 |
| Echino dermata | 0 | 24,1 | 0 | 0 | 4,7 | 0 | 0 |
| Mollusca | 0 | 7,9 | 0 | 0 | 3,6 | 0 | 0 |
| Polychaeta | 0 | 9,2 | 0 | 4,2 | 2,9 | 0 | 0 |
| Cod | 4,5 | 0,4 | 0,2 | 0 | 0,5 | 0,3 | 1,7 |
| Herring | 8,9 | 0,2 | 1,3 | 0,5 | 0,6 | 3,0 | 0 |
| Capelin | 11,6 | 2,1 | 8,7 | 30,8 | 17,5 | 54,9 | 0,9 |
| Haddock | 10,7 | 0,2 | 6,6 | 0,6 | 10,1 | 8,0 | 0 |
| Polar cod | 10,4 | 0 | 16,5 | 0 | 11,6 | 0,2 | 4,7 |
| Blue whiting | 4,8 | 0 | 2,6 | 0 | 0 | 0 | 0 |
| Greenland halibut | 0,2 | 0 | 1,4 | 0 | 0 | 0 | 0 |
| Redfish | 0,4 | 0 | 0,1 | 0 | 0 | 0 | 0 |
| Long rough dab | 1,8 | 0,1 | 4,8 | 2,9 | 0 | 0 | 0 |
| Other fish | 23,6 | 3,7 | 31,9 | 31,6 | 7,8 | 7,0 | 25,5 |
| Other food | 8,9 | 22,4 | 0,3 | 7,9 | 7,2 | 0 | 2,6 |
| Fishery waste | 0 | 4,1 | 17,7 | 4,9 | 31,4 | 0,9 | 0 |
| Undetermined | 0 | 2,4 | 0,2 | 1,4 | 0,7 | 0,5 | 0,3 |
| Total number of stomachs | 12209 | 7078 | 5223 | 432 | 2221 | 776 | 575 |
| Percentage of empty stomachs | 28,9 | 21,1 | 71,5 | 23,8 | 54,4 | 34,1 | 33,4 |
| Average filling degree | 1,7 | 1,6 | 0,7 | 1,9 | 1,1 | 1,6 | 1,7 |
| Mean index of stomach fullness | 213,8 | 110,5 | 84,4 | 182,7 | 139,0 | 116,3 | 111,2 |

Table 1.9. Annual consumption by minke whale and harp seal (thousand tonnes). The figures for minke whales are based on data from 1992-1995, while the figures for harp seals are based on data for 1990-1996.

| PREY | MINKE <br> CONSUMPTION | WHALE | HARP SEAL CONSUMPTION <br> (LOW CAPELIN STOCK) | HARP SEAL CONSUMPTION <br> (HIGH CAPELI STOCK) |
| :--- | :--- | :--- | :--- | :--- |
| Capelin | 142 | 23 | 812 |  |
| Herring | 633 | 394 | 213 |  |
| Cod | 256 | 298 | 101 |  |
| Haddock | 128 | 47 | 1 |  |
| Krill | 602 | 550 | 605 |  |
| Amphipods | 0 | 304 | $313^{2}$ |  |
| Shrimp | 0 | 1 | 1 |  |
| Polar cod | 1 | 880 | 608 |  |
| Other fish | 55 | 622 | 406 |  |
| Othercrustaceans | 0 | 356 | 312 |  |
| Total | 1817 | 3491 | 3371 |  |

${ }^{1}$ the prey species is included in the relevant 'other' group for this predator.
${ }^{2}$ only Parathemisto

Table 1.10. Description of the fisheries by gears. The gears are abbreviated as: trawl roundfish (TR), trawl shrimp (TS), longline (LL), gillnet (GN), handline (HL), purse seine (PS), Danish seine (DS) and trawl pelagic (TP). The regulations are abbreviated as: Quota (Q), mesh size (MS), sorting grid (SG), minimum catching size (MCS), maximum by-catch of undersized fish (MBU), maximum by-catch of non-target species (MBN), maximum as by-catch (MB), closure of areas (C), restrictions in season (RS), restrictions in area (RA), restriction in gear (RG), maximum by-catch per haul (MBH), as by-catch by maximum per boat at landing (MBL), number of effective fishing days (ED), number of vessels (EF).

| SPEC IES | DIRECTED FISHERY BY GEAR | TYPE OF FISH ERY | LANDINGS IN $2008^{\text {A }}$ (TONNES) | AS BY-CATCH IN FLEET(S) | LOCATION | AGREEMENTS AND REGULATIONS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Capelin | PS, TP | seasonal | $10000{ }^{\text {B }}$ | TR, TS | Northern coastal areas to south of $74{ }^{\circ} \mathrm{N}$ | Bilateral agreement, Norway and Russia |
| Coastal cod | $\begin{aligned} & \text { GN, LL, HL, } \\ & \text { DS } \end{aligned}$ | all year | 25 777C | $\begin{aligned} & \text { TS, PS, DS, } \\ & \text { TP } \end{aligned}$ | Norwegian coast (inside 12 naut.miles) north of $62^{\circ} \mathrm{N}$ | Q, MS, MCS, MBU, MBN, C, RS, RA |
| NEA Cod | TR, GN, LL, HL | all year | 464 171 ${ }^{\text {C }}$ | $\begin{aligned} & \mathrm{TS}, \mathrm{PS}, \mathrm{TP} \\ & \mathrm{DS} \end{aligned}$ | North of $62^{\circ} \mathrm{N}$, Barents Sea, Svalbard | Q, MS, SG, MCS, MBU, MBN, C, RS, RA |
| Wolffish | LL | all year | 11355 | TR, (GN), <br> (HL) | North of $62^{\circ} \mathrm{N}$, Barents Sea, Svalbard | Q, MB |
| Haddock | $\begin{aligned} & \text { TR, GN, LL, } \\ & \text { HL } \end{aligned}$ | all year | 155604 | $\begin{aligned} & \text { TS, PS, TP, } \\ & \text { DS } \end{aligned}$ | North of $62^{\circ} \mathrm{N}$, Barents Sea, Svalbard | Q, MS, SG, MCS, MBU, MBN, C, RS, RA |
| Saithe | PS, TR, GN | seasonal | 183443 | $\begin{aligned} & \text { TS, LL, HL, } \\ & \text { DS, TP } \end{aligned}$ | Coastal areas north of $62^{\circ} \mathrm{N}$, southern B arents Sea | Q, MS, SG, MCS, MBU, MBN, C, RS, RA |
| Greenland halibut | LL, GN | seasonal | 13144 | TR | Deep shelf and at the continental slope | Q, MS, RS, RG, MBH, MBL |
| Sebastes mentella | No directed fishery | all year | 13860 | TR | Pelagic in the Norwegian Sea, and as bycatch on the deep shelf and the continental slope | C, SG, MB |
| Sebastes <br> marinus | GN, LL, HL | all year | 6300 | TR | Norwegian coast and southwestern Barents Sea | SG, MB MCS, MBU, C |
| Shrimp | TS | all year | 21053 |  | Svalbard, <br> Barents Sea, Coastal north of $62^{\circ} \mathrm{N}$ | ED, EF, SG, C, MCS |

${ }^{\text {B }}$ On a research quota
${ }^{C}$ The total cod catch north of $62^{\circ} N(480,814 t)$ is the sum of the NEA cod catch given in the table above (464,171t) and the total cod catches between $62^{\circ} N$ and $67^{\circ} N$ for the whole year and between $67^{\circ} \mathrm{N}$ and $6^{\circ} \mathrm{N}$ for the second half of the year $(16,643 \mathrm{t})$.
${ }^{\mathrm{D}}$ The directed fishery for wolffish is mainly in ICES area IIb and the Russian EEZ, and the regulations are mainly restricted to this fishery
${ }^{\text {E }}$ Norwegian and Russian landings
${ }^{\text {F }}$ The only directed fishery for Greenland halibut is by a limited Norwegian fleet, comprising vessels less than 28 m .

Table 1.11. Flexibility in coupling between the fisheries. Fleets and impact on the other species (H, high, M, medium, L, low and 0 , nothing). The table below the diagonal indicates what gears couples the species, and the strength of the coupling is given above the diagonal. The gears are abbreviated as: trawl roundfish (TR), trawl shrimp (TS), longline (LL), gillnet (GN), handline (HL), purse seine (PS), Danish seine (DS) and trawl pelagic (TP).

| Species | Cod | Coastal cod | Haddock | Saithe | Wolffish | S. mentella | S. marinus | Greenland halibut | Capelin | Shrimp |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Cod |  | H | H | H | M | M | M | M | L | M-H <br> juvenile cod |
| Coastal cod | TR, PS,GN, LL, HL, DS |  | H | H | L | L | M-L | L | 0-L | L |
| Haddock | TR, PS, GN, LL, HL, DS | TR, PS, GN,LL, HL, DS |  | H | M | M | M | L | 0-L | M-H <br> juvenile haddock |
| Saithe | TR, PS, GN, LL, HL, DS | TR, PS, GN,LL, HL, DS | TR, PS, GN, LL, HL, DS |  | L | L | M | 0 | 0 | 0 |
| Wolffish | TR, GN, LL, HL | TR,GN, <br> LL, HL | TR, GN, LL, HL | TR, GN, LL, HL |  | M | M | M | 0 | M juvenile wolffish |
| S. mentella | TR | TR | TR | TR | TR |  | M | H | H <br> juvenile <br> Sebastes | H <br> juvenile <br> Sebastes |
| S. marinus | TR,GN, LL | TR,GN, LL | TR,GN, LL | TR,GN | TR, LL | TR |  | L | 0 | L-M juvenile Sebastes |
| Greenland halibut | TR, GN, LL,DS | TR,GN, LL | TR, GN, LL,DS | TR, GN, LL,DS | TR, LL | TR | TR |  | 0 | M-H <br> juvenile |
| Capelin | $\begin{aligned} & \text { TR, PS, TS, } \\ & \text { TP } \end{aligned}$ | PS, TP | $\begin{aligned} & \text { TR, PS, TS, } \\ & \text { TP } \end{aligned}$ | PS | TP | TP | TP | None |  | L |
| Shrimp | TS | TS | TS | TS | TS | TS | TS | TS | TS |  |

Table 1.12. The averaged annual relative changes (\%) of cod fishable stock biomass under various combinations of cod stock size, capelin stock size and water temperature according to STOCOBAR long-term simulations (the harvesting strategy is based on $\mathrm{Fpa}=0,5, \mathrm{Bpa}=460$ thousand tonnes). Different colours denotes different natural potential of cod to stock changes: red is high potential to stock decline, yellow is low potential to stock change, and red is high potential to stock increase.

| Temperature$\text { , } \mathrm{C}^{\circ}$ | Cod FSB* averaged for 3 previous years, millions t | Capelin stock biomass, millions t |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | <1 | 1-3 | 3-5 | >5 |
| <3,6 $\mathrm{C}^{\circ}$ | <1,4 | -8,96 | -1,73 | 4,90 | 5,60 |
|  | 1,4-1,8 | -11,60 | -7,28 | -3,89 | 2,98 |
|  | >1,8 | -16,89 | -12,56 | -7,94 | -3,61 |
| 3,6-4,2 $\mathrm{C}^{\circ}$ | <1,4 | 1,17 | 5,14 | 12,82 | 15,99 |
|  | 1,4-1,8 | -7,29 | -3,96 | 2,91 | 6,72 |
|  | >1,8 | -12,24 | -8,52 | -5,24 | -0,27 |
| $>4,2 \mathrm{C}^{\circ}$ | <1,4 | 3,77 | 7,14 | 16,78 | 20,94 |
|  | 1,4-1,8 | -2,53 | 1,36 | 8,34 | 16,96 |
|  | >1,8 | -3,62 | -0,95 | 1,25 | 1,31 |

*Fishable stock biomass

Table 1.13. Overview of available prognoses of NEA cod recruitment (in million individuals of age 3) from different models (sections 1.4.5) together with the 2009 assessment estimates (ICES AFWG 2009 Table 1.12).

| Model | Prognostic years | Updated | $\begin{aligned} & \hline 2010 \\ & \text { Prognoses } \end{aligned}$ | $\begin{aligned} & \hline 2011 \\ & \text { Prognoses } \end{aligned}$ | $\begin{aligned} & \hline 2012 \\ & \text { Progno ses } \end{aligned}$ | $2013$ <br> prognoses |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Titov0 | 0 | At assessment | 480 |  |  |  |
| Titov1 | $1\left(2^{1}\right)$ | At assessment | 518* | 470 |  |  |
| Titov2 | 2 | At assessment | 451 | $323 *$ |  |  |
| Titov3 | 3 | At assessment | 250* | 276* | 484* |  |
| Titov4 | 4 | At assessment | 425 | 362 | 780 | 946 |
| $\begin{aligned} & \hline \text { TB (1984- } \\ & 2000) \end{aligned}$ | 3 | Last year assessment | 632 | 553 |  |  |
| $\begin{aligned} & \hline \text { TB (1984- } \\ & 2004) \end{aligned}$ | 3 | Last year assessment | 627 | 551 |  |  |
| JES1 | $2\left(\begin{array}{ll} \\ \\ \end{array}\right)$ | At assessment | 878 | 797* | 827 |  |
| JES2 | $1\left(2{ }^{2}\right.$ | At assessment | 714 | 669 |  |  |
| JES3 | $0\left(\begin{array}{ll}1 \\ )\end{array}\right.$ | At assessment | 568 |  |  |  |
| H1 | 2 | At assessment | 890 | 889 |  |  |
| H2 | 2 | At assessment | 566 | 636 |  |  |
| H3 | 1 | At assessment | 500 |  |  |  |
| H4 | 1 | At assessment | 475 |  |  |  |
| $\begin{aligned} & \hline \text { RCT3 } \\ & 2010 \end{aligned}$ | 3 | At assessment | 289 | 558 | 675 |  |
| Hybrid <br> Model <br> (Assessment <br> 2009) |  | Last year assessment | 487 | 184 |  |  |
| Hybrid model <br> (Assessment 2010) |  | At assessment | 384 | 465 | 484 |  |

${ }^{1}$ Based on calculation of data from 2010.
${ }^{2}$ Based on prognosis estimate of capelin maturing biomass for October 1 2010, thereby allowing for an additional year.

* Models that are used in the Hybrid model at the 2010 assessment

Table 1.14.Proportion of cod in the diet of cod, based on Norwegian consumption calculations.

| Cod (predator)age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | $11+$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year |  |  |  |  |  |  |  |  |  |  |  |
| 1984 | 0.0000 | 0.0000 | 0.0032 | 0.0000 | 0.0437 | 0.0263 | 0.0328 | 0.0359 | 0.0367 | 0.0390 | 0.0374 |
| 1985 | 0.0015 | 0.0009 | 0.0014 | 0.0017 | 0.0314 | 0.0076 | 0.0827 | 0.0834 | 0.0842 | 0.0847 | 0.0853 |
| 1986 | 0.0000 | 0.0022 | 0.0015 | 0.0004 | 0.0130 | 0.1761 | 0.1767 | 0.1766 | 0.1762 | 0.1757 | 0.1748 |
| 1987 | 0.0000 | 0.0000 | 0.0007 | 0.0051 | 0.0103 | 0.0246 | 0.0377 | 0.0400 | 0.0418 | 0.0405 | 0.0436 |
| 1988 | 0.0000 | 0.0000 | 0.0000 | 0.0002 | 0.0058 | 0.0014 | 0.0038 | 0.0036 | 0.0032 | 0.0038 | 0.0036 |
| 1989 | 0.0000 | 0.0006 | 0.0016 | 0.0019 | 0.0027 | 0.0040 | 0.0035 | 0.0035 | 0.0040 | 0.0038 | 0.0041 |
| 1990 | 0.0000 | 0.0000 | 0.0000 | 0.0012 | 0.0017 | 0.0019 | 0.0268 | 0.0268 | 0.0268 | 0.0268 | 0.0268 |
| 1991 | 0.0000 | 0.0005 | 0.0000 | 0.0003 | 0.0032 | 0.0020 | 0.0224 | 0.0232 | 0.0235 | 0.0239 | 0.0241 |
| 1992 | 0.0000 | 0.0021 | 0.0037 | 0.0129 | 0.0250 | 0.0475 | 0.0120 | 0.0159 | 0.0232 | 0.0232 | 0.0230 |
| 1993 | 0.0000 | 0.0413 | 0.0368 | 0.0515 | 0.0536 | 0.1156 | 0.0498 | 0.0801 | 0.0801 | 0.0801 | 0.0805 |
| 1994 | 0.0000 | 0.0038 | 0.0917 | 0.0347 | 0.0285 | 0.0784 | 0.1247 | 0.1339 | 0.2617 | 0.2634 | 0.2608 |
| 1995 | 0.0069 | 0.0811 | 0.0745 | 0.0802 | 0.0925 | 0.1123 | 0.1389 | 0.2533 | 0.2553 | 0.2561 | 0.2574 |
| 1996 | 0.0000 | 0.1493 | 0.2549 | 0.2060 | 0.1322 | 0.1267 | 0.1850 | 0.2082 | 0.2459 | 0.2471 | 0.2465 |
| 1997 | 0.0000 | 0.0704 | 0.0767 | 0.1140 | 0.1552 | 0.1554 | 0.2329 | 0.2267 | 0.2882 | 0.2815 | 0.2832 |
| 1998 | 0.0000 | 0.0135 | 0.0272 | 0.0418 | 0.1041 | 0.0981 | 0.1081 | 0.1492 | 0.2758 | 0.2767 | 0.2778 |
| 1999 | 0.0000 | 0.0000 | 0.0049 | 0.0137 | 0.0148 | 0.0338 | 0.0620 | 0.1117 | 0.1937 | 0.1940 | 0.1840 |
| 2000 | 0.0000 | 0.0000 | 0.0286 | 0.0147 | 0.0134 | 0.0266 | 0.0499 | 0.0566 | 0.2757 | 0.2726 | 0.2738 |
| 2001 | 0.0000 | 0.0158 | 0.0116 | 0.0082 | 0.0131 | 0.0241 | 0.0496 | 0.0381 | 0.3294 | 0.3264 | 0.3301 |
| 2002 | 0.0000 | 0.0387 | 0.0591 | 0.0142 | 0.0187 | 0.0285 | 0.0359 | 0.0627 | 0.1603 | 0.1575 | 0.1581 |
| 2003 | 0.0000 | 0.0193 | 0.0198 | 0.0199 | 0.0206 | 0.0188 | 0.0456 | 0.1043 | 0.2257 | 0.2281 | 0.2269 |
| 2004 | 0.0230 | 0.0223 | 0.0294 | 0.0214 | 0.0184 | 0.0294 | 0.0391 | 0.0710 | 0.1059 | 0.1056 | 0.1061 |
| 2005 | 0.0000 | 0.0261 | 0.0229 | 0.0258 | 0.0155 | 0.0241 | 0.0487 | 0.0830 | 0.1688 | 0.1667 | 0.1693 |
| 2006 | 0.0000 | 0.0051 | 0.0007 | 0.0130 | 0.0285 | 0.0124 | 0.0397 | 0.0316 | 0.0841 | 0.0845 | 0.0834 |
| 2007 | 0.0000 | 0.0000 | 0.0010 | 0.0108 | 0.0137 | 0.0314 | 0.0336 | 0.0724 | 0.1518 | 0.1543 | 0.1504 |
| 2008 | 0.0000 | 0.0821 | 0.0243 | 0.0068 | 0.0089 | 0.0110 | 0.0820 | 0.1004 | 0.1223 | 0.1212 | 0.1198 |
| 2009 | 0.0376 | 0.0353 | 0.0227 | 0.0137 | 0.0147 | 0.0250 | 0.0981 | 0.0918 | 0.0920 | 0.0919 |  |
|  | 0.0312 | 0.0278 | 0.0339 | 0.0474 | 0.0673 | 0.0881 | 0.1437 | 0.1434 | 0.1432 |  |  |

Table 1.15. Cannibalism mortality in cod.

| Year | M at age 3 | M at age 4 |
| :---: | :---: | :---: |
|  | by regression |  |
| 2009 | 0.40 | 0.27 |
| 2010 | 0.43 | 0.28 |
| 2011 | 0.46 | 0.29 |
| 2012 | 0.64 | 0.36 |
|  | values used in assessment |  |
| 2010-2012 | 0.3335 | 0.227 |

Table 1.16. Overview of the standard sections monitored by IMR and P INRO in the Barents Sea, with observed parameters. Parameters are: T-temperature, S-Salinity, N -nutrients, chlachlorophyll, zoo-zooplankton, O-oxygen.

| SECTION | Institution | TIME PERIOD | Ob servation FREQUENCY | PARAMETERS |
| :---: | :---: | :---: | :---: | :---: |
| Fuglay a-Bear Island | IMR | 1977-present | 6 times pryear | T,S,N,chla,zoo |
| North cape-Bear Island | PINRO | 1950's-present | yearly | T, S |
| Bear Island-East | PINRO | 1950's-present | yearly | T, S |
| V ardø-North | IMR | 1977-present | 4 times pr year | T,S,N,chla |
| Kola | PINRO | 1921-present | monthly | T,S,O,N |
| Kanin | PINRO | 1950's-present | yearly | T, S |
| Sem Islands | IMR | 1970's-present | Intermittently* | T, S |

* The Sem Island section is not observed each year, and have not been observed the last 3-4 years.

Table 1.17. Overview of conducted monitoring surveys by IMR and PINRO in the Barents Sea, with observed parameters and species. For zooplankton, mammals and benthos abundance and distribution for many species are investigated. Therefore, in the table it is only indicated whether sampling is conducted. Climate and phytoplankton parameters are: T-temperature, S-Salinity, N-nutrients, chla-chlorophyll.

| SURVEY | INSTITUTION | PERIOD | Climate | PHYTO- <br> PLANKTON | ZOO-PLANKTON | JUVENILE FISH | $\begin{aligned} & \text { TARGET FISH } \\ & \text { STOCKS } \end{aligned}$ | MAMMALS | Benthos |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Winter | Joint | Feb-Mar | T,S | N, chla | intermittent | All commercial species and some additional | Cod, Haddock | - | - |
| Lofoten | IMR | Mar-Apr | T,S | - | - |  | Cod, haddock, saithe | - | - |
| Ecosystem survey | Joint | Aug-Oct | T,S | N,chla | Yes | All commercial species and some additional | All commercial species and some additional | Yes | Yes |
| Norwegian coastal surveys | IMR | Oct-Nov | T,S | N,chla | Yes | Herring, sprat, demersal species | Saithe, coastal cod | - | - |
| Autumnwinter trawlacoustic survey | PINRO | Oct-Des | T,S | - | Yes | Demersal species | Demersial species | - | - |
| Norwegian Greenland halibut survey | IMR | Aug | - | - | - | - | Greenland halibut, redfish | - | - |
| Russian young herring survey | PINRO | May | T, S |  | Yes |  | Herring | - | - |



Figure 1.1. The main features of the circulation and bathymetry of the Barents Sea.


Figure 1.2. Temperature (upper) and salinity (lower) anomalies in the $50-200 \mathrm{~m}$ layer of the Fugloya-Bear Island section (left) and the Vardø-North section (right).


Figure 1.3. Monthly temperature anomalies in the 0-200 m layer of the Kola Section in 2008 (black dots) and 2009 (red bars). St. 1-3 - Coastal waters, St. 3-7 - Murman Current (Anon., 2010).


Figure 1.4. Bottom temperature anomalies in the Barents Sea in August-September 2009 (Anon., 2010).


Figure 1.5. Observed Atlantic Water volume flux through the Fugløya-Bear Island section estimated from current meter moorings. Three months (blue line) and 12 -months (red line) running means are shown.


Figure 1.6. Ice extent anomalies in the Barents Sea in 1982-2009 (Anon., 2010). The blue line shows monthly values, the red one - 11-month moving average values.


Figure 1.7. Horizontal distribution of zooplankton in 2009 ( $\mathrm{g} \mathrm{m}^{-2}$ of dry weight from bottom- 0 m ).


Figure 1.8. The mean abundance of euphausiids in the north-western and western areas of the Barents Sea in 2008 and 2009.


Figure 1.9. Biomass of pelagic fish species in the Barents Sea. Data are taken from; capelin: Acoustic estimates in September, age 1+ (ICES AFWG 2010), herring: VPA estimates of age 1 and 2 herring (ICES C.M. 2009/ACOM:12), using standard weights at age ( 9 g for age 1 and 20 g for age 2); polar cod and blue whiting: Acoustic estimates in September, age 1+ (Anon. 2010), 0-group: estimates of biomass of cod, haddock, herring and capelin 0 -group, corrected for catching efficiency (Eriksen et al. 2010).


Figure 1.10. Biomass of demersal fish species in the Barents Sea. Data are taken from; cod: VPA estimates, age 3+ (ICES AFWG 2010); haddock: VPA estimates, age 3+ (ICES, 2010); Greenland halibut: VPA estimates, age 5+ (ICES, 2007); Sebastes mentella: VPA estimates, age 6+ (ICES, 1995 for the years 1968-1990; ICES, 2003 for the years 1991-2002).


Figure 1.11a. Time series of annual average fishing mortalities for Northeast Arctic cod (time period 1946-2009, average for ages 5-10), Northeast Arctic saithe (time period 1960-2009, average for ages 4-7), coastal cod (1984-2009, average for ages 4-7), Northeast Arctic haddock (time period 1950-2009, average for ages 4-7), Greenland halibut (time period 1964-2009, average for ages 6-10) and Sebastes marinus (time period 1990-2009, average for ages 12-\#).


Figure 1.11b. Left panel - annual fishing mortalities of the Northeast Arctic cod, haddock and saithe stocks relative to the critical levels above which the fishing mortality will impair the recruitment. Right panel - annual fishing mortalities of Golden redfish (Sebastes marinus) and Greenland halibut (Reinhardtius hippoglossoides) relative to the proposed maximum levels above which the fishing mortality over time most probably will impair the recruitment.


Figure 1.12. Pair-wise plots of annual ave rage fishing mortalities (above diagonal) and landings (below diagonal) for ove rlapping time periods for Northeast Arctic cod (time period 1946-2009, ave rage for ages 5-10), Northeast Arctic haddock (time period 1950-2009, ave rage for ages 4-7), Northeast Arctic saithe (time period 1960-2009, ave rage for ages 4-7), coastal cod (1984-2009, ave rage for ages 4-7), Greenland halibut (time period 1964-2009, ave rage for ages 6-10) and Sebastes marinus (time period 1990-2009, ave rage for ages 12-19). The correlation and the corresponding pvalue are given in the legend.


Figure 1.13. Relative distribution by weight of cod, haddock, saithe, Greenland halibut, golden redfish (Sebastes marinus), beaked redfish (Sebastes mentella) and other species taken by Russian bottom trawl in 2009 per main area for the Russian strata system.


Figure 1.14. Relative distribution by weight of Norwegian catches of cod, haddock, and saithe per main area for the Norwegian strata system.


Figure 1.15. The Russian catch of cod, haddock, saithe, Greenland halibut, Sebastes marinus, Sebastes mentella and other species taken by bottom trawl by main statistical areas in 2009, thousand tonnes. The statistical areas correspond to the areas shown in Figure 1.13.



Figure 1.16. The Norwegian catch of cod, haddock and saithe by main statistical areas in 2009, thousand tonnes. The statistical areas correspond to the areas shown in Figure 1.14.


Figure 1.17. Upper panel - gear composition of the Norwegian groundfish (2007; left) and pelagic capelin (2000-2008; right) fisheries in the Northeast Arctic. Note that the purse seine in the groundfish fishery is solely used in a coastal fishery for saithe. Lower panel - gear composition of the Russian groundfish (2007; left) and pelagic capelin (2000-2008; right) fisheries in the Northeast Arctic.


Figure 1.18. Positions of the standard sections monitored in the Barents Sea. $A$ is fixed station Ingøy, B is Fugløya-Bear Island, C is North cape-Bear Island, D is Vardø-North, E is Kola, F is Sem Island-North G is Kanin section and H is Bear Island-East section.

## 1 Ecosystem considerations (Figures 1.1-1.18, Tables 1.1-1.17)

The aim of this chapter is to identify important ecosystem information influencing the fish stocks, and further show how this knowledge may be implemented into the fish stock assessment and predictions. There has been a steadily development in this aspect over the last few years and the work is still in a developing phase. Hopefully, the gathering of information on the ecosystem in this chapter will lead to a better understanding of the complex dynamics and interactions that takes place in the ecosystem, and also participate in the development of an ecosystem based management of the Barents Sea.

The ecosystem approach to management is variously defined, but in principle it puts emphasis on a management regime that maintains the health of the ecosystem alongside appropriate use of the marine environment, for the benefit of current and future generations (Jennings, 2004).

Along with fishery, changes in the Barents Sea ecosystem are mainly caused by variations in the ocean climate. A warm period is characterized of increased impact of warm Atlantic water in the Barents Sea contributes to advection of zooplankton, faster growth rate in fish and emergence of abundant year classes (Dalpadado et al. 2002). A cold period is, conversely, characterized by reduced primary biological production in the Barents Sea and emergence of weak year classes of commercial species. Climatic conditions govern the formation of primary biological production and feeding conditions for fish, as well as the survival of their offspring. In addition, inter-species trophic relations are an important factor that influences the abundance dynamics of commercial species.

Movement towards an ecosystem approach to the fishery management in the Barents Sea should include (Filin and Røttingen, 2005):

- More extensive use of ecosystem information in the population parameters applied in assessment and prognosis,
- Expansion of the use of multi-species models for fishing management.

This chapter is in general based on a preliminary version of the 2009 update (Stiansen et al., WD23) of the "Joint Norwegian-Russian environmental statutes 2008, report on the Barents Sea Ecosystem" (Stiansen et al., 2009), affiliating more than 100 scientists from 24 institutions in Norway and Russia. This report is the successor to the "Joint PINRO/IMR report on the state of the Barents Sea ecosystem in 2007, with expected situation and considerations for management" (Stiansen and Filin, 2008). Text, figures and tables taken from these reports (i.e. Stiansen et al., 2009, or Stiansen et al, WD23) are in general not further cited in this chapter.

### 1.1 General description of the Barents Sea ecosystem (Figure 1.1, Tables

## 1.1-1.7)

## Geographical description

The Barents Sea is on the continental shelf surrounding the Arctic Ocean. It connects with the Norwegian Sea to the west and the Arctic Ocean to the north. Its contours are delineated by the continental slope between Norway and Spitsbergen to the west, the top of the continental slope towards the Arctic Ocean to the north, Novaya Zemlya archipelago to the east, and the coasts of both Norway and Russia to the south
(Figure 1.1). It covers an area of approximately 1.4 million $\mathrm{km}^{2}$, has an average depth of 230 m , and a maximum depth of about 500 m at the western end of Bear Island Trough (Figure 1.1). Its topography is characterized by troughs and basins ( $300 \mathrm{~m}-$ 500 m deep), separated by shallow bank areas, with depths ranging from 100-200 m. The three largest banks are Central Bank, Great Bank and Spitsbergen Bank. Several troughs over 300 m deep run from central Barents Sea to the northern (e.g. Franz Victoria Trough) and western (e.g. Bear Island Trough) continental shelf break. These troughs allow the influx of Atlantic waters to the central Barents Sea.

## Climate

The general pattern of circulation (Figure 1.1) is strongly influenced by this topography, and is characterised by inflow of relatively warm Atlantic water, and coastal water from the west. This current divides into two branches: 1) a southern branch that flows parallel to the coast and eastwards towards Novaya Zemlya; and 2) a northern branch that flows into the Hopen Trench. The Coastal Water has more freshwater runoff and a lower salinity than the Atlantic water; it also has a stronger seasonal temperature signal. In the northern region of the Barents Sea, fresh and cold Arctic waters flow from northeast to southwest. Atlantic and Arctic water masses are separated by the Polar Front, which is characterised by strong gradients in both temperature and salinity. There is large inter-annual variability in ocean climate related to variable strength of the Atlantic water inflow, and exchange of cold Arctic water. Thus, seasonal variations in hydrographic conditions can be quite large.

## Bacteria and phytoplankton

In the biogeochemical cycles of the ocean, a multitude of processes are catalyzed by Bacteria and Archaea, and the functioning of these cycles in the Barents sea do not differ qualitatively from those at lower latitudes. Both bacteria and viruses show highly variable abundance in the Barents Sea, and in general, the dynamics of these groups in this area do not differ from other parts of the ocean. The situation in the icecovered areas in the north remains to be investigated.

The Barents Sea is a spring bloom system. During winter, primary production is close to zero. Timing of the phytoplankton bloom varies throughout the Barents Sea and there may also be a high inter-annual variability. The spring bloom starts in the south-western areas and spreads north and east with the retracting ice. In early spring, the water is mixed from surface to bottom. Despite adequate nutrient and light conditions for production, the main bloom does not occur until the water becomes stratified.

Stratification of water masses in different areas of the Barents Sea may occur in several different ways; 1) through fresh surface water from melting ice along the marginal ice zone; 2) through solar heating of surface layers in Atlantic water masses; or 3) through lateral dispersion of waters in the southern coastal region (Rey, 1981). As in other areas, diatoms are also the dominant phytoplankton groups in the Barents Sea (Rey, 1993). Diatoms particularly dominate the first part of the spring bloom, and the concentration of diatoms can reach up to several million cells per litre. They require silicate for growing, and when this is consumed, other phytoplankton groups, such as flagellates, take over. An important flagellate species in the Barents Sea is Phaeocystis pouchetii but other species may, however, predominate the spring bloom in different years.

## Zooplankton

In the Barents Sea ecosystem, zooplankton forms a link between phytoplankton (primary producers) and fish, mammals and other organisms at higher trophic levels. Zooplankton biomass in the Barents Sea can vary significantly between years and crustaceans are important. The calanoid copepods of the genus Calanus play a key role in this ecosystem. Calanus finmarchicus, is most abundant in Atlantic waters and C. glacialis is most abundant in Arctic waters. Both form the largest component of zooplankton biomass.

Calanoid copepods are largely herbivorous, and feed particularly on diatoms (Mauchline, 1998). Krill (euphausiids), another group of crustaceans, also play a significant role in the Barents Sea ecosystem as food for fish, seabirds, and marine mammals. Krill species are believed to be omnivorous: filter-feeding on phytoplankton during the spring bloom; while feeding on small zooplankton during other times of the year (Melle et al., 2004). Four dominant species that occupy different niches in the community of Barents Sea euphausiids are: Meganyctiphanes norvegica (neritic shelf boreal); Thysanoessa longicaudata (oceanic arcto-boreal); T. inermis (neritic shelf arcto-boreal); and T. raschii (neritic coastal arcto-boreal) (Drobysheva, 1994). The two latter species comprise $80-98 \%$ of total euphausiid abundance, but species composition may vary between years relative to climate (Drobysheva, 1994). After periods with cold climate, observed abundance of $T$. raschii increased while abundance of $T$. inermis decreased (Drobysheva, 1967). Advection from the Norwegian Sea is influenced by the intensity of Atlantic water inflow, which also influences the composition of species (Drobysheva, 1967; Drobysheva et al., 2003).

Three amphipod species were found abundant in the Barents Sea; Themisto abyssorum and T. libellula in the western and central Barents Sea, and T. compressa is found, albeit less abundant, in central and northern regions. T. abyssorum is most abundant in subArctic waters. In contrast, the largest of the Themisto species, T. libellula, is largely restricted to combined Atlantic and Arctic water masses. High abundance of T. libellula was observed adjacent to the Polar Front. Amphipods feed on small zooplankton and copepods form an important component of their diet (Melle et al., 2004).
"Gelatinous zooplankton" is a term often used by non-specialists in reference to classes of organism that are jelly-like in appearance. The term "jellyfish" is commonly used in reference to marine invertebrates belonging to the class Scyphozoa, phylum Cnidaria. Neither of these terms implies any systematic relationship to vertebrate fish. The term "jellyfish" is also often used in reference to relatives of true scyphozoans, particularly the Hydrozoa and the Cubozoa. Both comb-jellies (Ctenophora) and "true" jellyfish are predators, and they compete with plankton-eating fish, because copepods often are significant prey items.

## Benthos

The sea floor is inhabited by a wide range of organisms. Some are buried in sediment, others are attached to a substrate, some are slow and sluggish, others roving and rapid. Many feed by actively or passively, sieving food particles or small organisms from the water. Others eat the bottom sediments (detritus feeders), eat carrion (scavengers) or hunt other animals (carnivores). The high diversity among bottom animals is presumed to be due to the abundance of micro-habitats that organisms can adapt. In shallow waters, kelp forests are feeding and nursery habitats for many species of fish, birds, and mammals. Below the sublittoral zone, sea anemones, sponges, hydrozoans, tunicates, echinoderms, crustaceans, molluscs and many other animal groups
abound on hard substrates. These large conspicuous animals are not abundant on sand or muddy bottoms, and in fact some of these habitats may at first look rather lifeless. However, most of the benthic animals in these habitats live buried in the sediments. Polychaete worms, crustaceans and bivalves are found in the sediments well as a myriad of other taxa. Some muddy areas might have dense aggregations of brittle stars, sea stars or bivalves.

More than 3050 species of benthic invertebrates inhabit the Barents Sea (Sirenko, 2001). The benthic ecosystems in the Barents Sea have considerable value, both in direct economic terms, and in their ecosystem functions. Scallops, shrimp, king crab, and snow crab are benthic residents which are harvested in the region. Many species of benthos are also interesting for bio-prospecting or as a future food resource, such as sea cucumber, snails and bivalves. Several of them are crucial to the ecosystem. Important fish species such as haddock, catfish and most flatfishes primarily feed on benthos. Many benthic animals, primarily bivalves, filter particles from the ocean and effectively clean it up. Others scavenge on dead organisms, returning valuable nutrients to the water column. Detritus feeders and other active diggers regularly move the bottom sediments around and therefore increase sediment oxygen content and overall productivity - much like earthworms on land.

The northern shrimp (Pandalus borealis) is distributed in most deep areas of the Barents Sea and Spitsbergen waters. The densest concentrations are found in depths between 200 and 350 meter. The shrimp mainly feed on detritus, but may also be a scavenger. Shrimp is also important as a food item for many fish species and seals.

Red king crab (Paralithodes camtschatica) was introduced to the Barents Sea in the 1960s. Presently it is an important commercial species. Adult red king crabs are opportunistic omnivores.
The snow crab (Chionoecetes opilio) is an invasive species in the Barents Sea. The first recordings of this species in the Barents Sea were in 1996. Since 2003 snow crab have been found in the stomachs of cod, haddock, catfishes and thorny skates that indicates that the crab abundance and settlement density substantially increased.

## Fish

More than 200 fish species are registered in trawl catches during surveys of the Barents Sea, and nearly 100 of them occur regularly. Even so, the Barents Sea is a relatively simple ecosystem, with few fish species of potentially high abundance. Different species of fish are not evenly distributed throughout the Barents Sea. Rather, they exhibit highest abundance in areas with suitable environmental conditions. Commercially important fish species include Northeast Arctic cod, Northeast Arctic haddock, Barents Sea capelin, polar cod and immature Norwegian springspawning herring. In warm years, increased numbers of young blue whiting have migrated into the Barents Sea. Species distribution largely depends on positioning of the Polar Front. There have been significant variations in abundance of these species. These variations are due to a combination of fishing pressure and environmental variability Cod, capelin, and herring are key species in the Barents Sea trophic system.

In general the four pelagic species (herring, capelin, polar cod and blue whiting) have minor overlapping distributions; with the blue whiting in the west, the herring in the south, the polar cod in the east (except for an overlapping part of the stock in the Svalbard region) and the capelin in the north and central areas. In southwestern areas blue whiting and herring partly overlap. However, they occupy different parts of the water column.

The recruitment of the Barents Sea fish species has shown a large year-to-year variability (Tables 1.1-1.2). The most important reasons for this variability are variations in the spawning biomass, climate conditions, food availability and predator abundance and distribution. Variation in the recruitment of some species, like cod, haddock and herring, has been associated with changes in the influx of Atlantic waters into the Barents Sea.

Cod prey on capelin, herring, and smaller cod; while herring prey on capelin larvae. Cod is the most important predator fish species in the Barents Sea, and feeds on a wide range of prey, including larger zooplankton, most available fish species and shrimp (Table 1.3-

Table 1.6). Cod prefer capelin as a prey, and fluctuations of the capelin stock (Table 1.7) have a strong effect on growth, maturation and fecundity of cod, as well as on cod recruitment because of cannibalism. The role of euphausiids for cod feeding increases in the years when capelin stock is at a low level (Ponomarenko and Yaragina $1990)$. Also, according to Ponomarenko $(1973,1984)$ interannual changes of euphausiid abundance is important for the survival rate of cod during the first year of life.

Capelin feed on zooplankton produced near the ice edge. Farther south, capelin is the most important prey species in the Barents Sea as it transports biomass from northern to southern regions (von Quillfeldt and Dommasnes, 2005). The capelin has showed large variations in abundance.

Herring is also a major predator on zooplankton. The herring spawns along the Norwegian western coast and the larvae drifts into the Barents Sea. The juveniles of the Norwegian spring-spawning herring stock are distributed in the southern parts of the Barents Sea. They stay in this area for about three years before they migrate west and southwards along the Norwegian coast and mix with the adult part of the stock. The presence of young herring in the area has a profound effect on the recruitment of capelin, and it has been shown that when rich year classes of herring enters to the Barents Sea, the recruitment to the capelin stock is poor, and in the following years the capelin stock collapses (Gjøsæter and Bogstad, 1998).

Haddock is also a common species, and migrates partly out of the Barents Sea. The stock has large natural variations in stock size. Food composition of haddock consists mainly of benthic organisms.

Saithe is found mainly along the Norwegian coast, but also occurs in the Norwegian Sea and in the southern Barents Sea. The 0-group saithe drifts from the spawning grounds to inshore waters. The smaller individuals feed on crustaceans, while larger saithe depends more on fish as prey (Dolgov, WD 29, AFWG 2006; Mehl, WD7, AFWG 2005). The main fish prey is young herring, Norway pout, haddock, blue whiting and capelin, while the dominating crustacean prey is krill. Polar cod is a cold-water species found particularly in the eastern Barents Sea and in the north. It seems to be an important forage fish for several marine mammals, but to some extent also for cod. There is little fishing on this stock.

Deep-sea redfish and golden redfish used to be important elements in the fish fauna in the Barents Sea, but due to heavy overfishing these stocks declined strongly during the 1980's, and has since then stayed at a low level. Young redfish are plankton eaters, but larger individuals take larger prey, including fish.

Greenland halibut is a large and voracious fish predator with the continental slope between the Barents Sea and the Norwegian Sea as its most important area, but it is also found in the deeper parts of the Barents Sea. Investigations in the period 1980-

1990 showed that cephalopods (squids, octopuses) dominated in the Greenland halibut stomachs, as well as fish, mainly capelin and herring. Ontogenetic shift in prey preference was clear with decreasing proportion of small prey (shrimps and small capelin) and increasing proportion of larger fish with increasing predator length. The largest Greenland halibut (length more than 65-70 cm) had a rather big portion of cod and haddock in the diet.

The blue whiting has its main distribution area in the Norwegian Sea and Northeast Atlantic, and the marginal northern distribution is at the entrance to the Barents Sea. Usually the blue whiting population in the Barents Sea is small. In years with warm Atlantic water masses the blue whiting may enter the Barents Sea in large numbers, and the blue whiting can be a dominant species in the western areas. This situation occurred from 2001 onwards, and blue whiting were found in great numbers for the period 2003-2007. Since then it has decreased strongly again. This rise and fall is probably due to a combination of variation in stock size and environmental conditions. In the diet of blue whiting zooplankton(copepods, hyperiids and euphausiids) is dominant in the younger age groups, while fish is increasingly important as the blue whiting gets older(Dolgov, WD 29, AFWG 2006).

Long rough dab is a typical ichthyobenthophage, which mainly eats benthos (ophiura, polychaetes etc.) and different fish species (Dolgov, WD 29, AFWG 2006). At older stages the proportion of fish in the diet increases (polar cod and cod, capelin and juvenile redfish). The larger long rough dab also feed on their own juveniles and juvenile haddock.

Thorny skate preys primarily on large crustaceans, shrimps and crabs (Dolgov, WD 29, AFWG 2006), but may also in a lesser extent feed on fish. The most common fish species are young cod and capelin. Round skate fed mainly on benthos, especially Polychaeta and Gammaridae. Arctic skate feed mainly on fish and shrimp (herring, capelin, redfish and northern shrimp). Blue skate diet consists largely of fish, mainly young cod and haddock, redfish, and long rough dab). Spinytail skate also prey mostly on fish, which included haddock, redfish and long rough dab. Total yearly food consumption by thorny skate is estimated to be around 160 thousand tonnes, of which around 75 thousand tonnes comprised commercial fishes and invertebrates. Total yearly food consumption by all other skate species was estimated to be around 30 thousand tonnes, of which around 20 thousand tonnes was commercial species (Dolgov, WD 29, AFWG 2006).

## Mammals and seabirds

Marine mammals, as top predators, are keystone species significant components of the Barents Sea ecosystem. About 25 species of marine mammals regularly occur in the Barents Sea, including: 7 pinnipeds (seals and walruses); 12 large cetaceans (large whales); 5 small cetaceans (porpoises and dolphins); and the polar bear (Ursus maritimus). Some of these species are not full-time residents in the Barents Sea, and use temperate areas for mating, calving, and feeding (e.g. minke whale Balaenoptera acutorostrata). Others reside in the Barents Sea all year round (e.g. white-beaked dolphin Lagenorhynchus albirostris, and harbour porpoise Phocoena phocoena). Some marine mammals are naturally rare, such as the beluga whale Delphinapterus leucas. Others are rare due to historic high exploitation, such as bowhead whale Balaena mysticetus and blue whale Balaenoptera musculus.

Marine mammals may consume up to 1.5 times the amount of fish caught in fisheries. Minke whales and harp seals may each year consume 1.8 million and 3-5 million tons
of prey of crustaceans, capelin, herring, polar cod, and gadoid fish respectively (Folkow et al., 2000; Nilssen et al., 2000). Functional relationships between marine mammals and their prey seem closely related to fluctuations in marine ecosystems. Both minke whales and harp seals are thought to switch between krill, capelin and herring depending on availability of the different prey species (Lindstrøm et al., 1998; Haug et al., 1995; Nilssen et al., 2000).

Fish and mammals have seasonal feeding migrations so that the stocks in the area will have their most northern and eastern distribution in August-September and be concentrated in the southern and south-western areas in February-March. The Barents Sea has one of the largest concentrations of seabirds in the world (Norderhaug et al., 1977; Anker-Nilssen et al., 2000); its 20 million seabirds harvest annually approximately 1.2 million tonnes of biomass from the area (Barrett et al., 2002). Nearly 40 species are thought to breed regularly in northern regions of the Norwegian Sea and the Barents Sea. Abundant species belong to the auk and gull families. Seabirds play an important role in transporting organic matter and nutrients from the sea to the land (Ellis, 2005). This transport is of great importance especially in the Arctic, where lack of nutrients is an important limiting factor.

## Rare, threatened and invasive species and infectious organisms

There are 10 types of parasites found in the fish of the Barents Sea, but it is hard to determine which groups of parasitic organisms that play an important role in the population dynamics of their hosts. The Barents Sea parasites considered to be most damaging to the human health are larvae stages of Cestoda (Diphyllobothrium and Pyramicocephalus genera), Nematoda (Anisakis and Pseudoterranova genera) and Palaeacanthocephala (Corynosoma genera). 82 species of helminthes are recorded from 18 bird species. The Barents Sea birds' helminthofauna mostly consists of the species with the life cycle dependent on coastal ecosystems. Invertebrates and fish from the littoral and upper sub littoral complex serve as their intermediate hosts.

The Barents Sea includes species that either have very small populations or species that have recently undergone considerable population decline (or are expected to do so in the close future). The assessments are done by use of the IUCN criteria (IUCN, 2001; 2003), but the Global, the Russian and the Norwegian lists available cannot be directly compared. All these lists are closely related and have high relevance for the conservation of biodiversity, and the list from the Barents Sea include a total of 56 species comprising of 28 fish species, 9 bird species, and 18 mammal species.

Invasions of alien species - spread of the representatives of various groups of living organisms beyond their primary habitats - are global in nature. Their introduction and further spread often leads to the undesirable environmental, economic and social consequences. Different modes of biological invasions can be natural movement associated with the population dynamics and climatic changes, intentional introduction and reintroduction, and accidental introduction with the ballast waters and along with the intentionally introduced species, etc. The best known examples of introduced species in the Barents Sea are red king crab (Paralithodes camtschaticus) and snow crab (Chionoecetes opilio).

## Human activity

The Barents Sea is strongly influenced by human activity; historically involving the fishing and hunting of marine mammals. More recently, human activities also involve transportation of goods, oil and gas, tourism and aquaculture. In the last years
interest has increases on the evaluation of the most likely response of the Barents Sea ecosystem to the future climate changes due to anthropogenic effects on climate warming.

Fishing is the largest human impact to the fish stocks in the Barents Sea, and thereby the functioning of the whole ecosystem. However, the observed variation in both fish species and ecosystem is also impacted by other effects such as climate and predation. The most widespread gear used in the central Barents Sea is bottom trawl, but also long line and gillnets are used in the demersal fisheries. The pelagic fisheries use purse seine and pelagic trawl.

The Barents Sea remains relatively clean, however, when compared to marine areas in many industrialized parts of the world. Major sources of contaminants in the Barents Sea are natural processes, long-range transport, accidental releases from local activities, and ship fuel emissions. Results of recent studies indicate low level of contaminants in the Barents Sea marine environment and confirm results of earlier studies on bottom sediments in the same areas. In the near-term, observed levels of contaminants in the marine environment should not have significant impact on commercially important stocks and on the Barents Sea ecosystem as a whole.

Traditionally, fishing having been the most important and far-reaching human activity in the ecosystem has been given most of the attention with analyses of impacts and risks. This need has increased in importance as oil- and gas industries have begun to develop new off-shore fields in the Barents Sea, and ship transport of oil and gas from the region has increased exponentially over the last 5 years.

The Barents Sea can become an important region for oil and gas development. Currently offshore development is limited both in the Russian and Norwegian economic zones (to the Snøhvit field north of Hammerfest in the Norwegian zone), but this may increase in the future with development of new oil- and gas fields. In Russia there are plans for the development of Stockman, a large gas-field west of Novaya Zemlja. The environmental risk of oil and gas development in the region has been evaluated several times, and is a key environmental question facing the region.

Transport of oil and other petroleum products from ports and terminals in NWRussia have been increasing over the last decade. In 2002, about 4 million tons of Russian oil was exported along the Norwegian coastline, in 2004, the volume reached almost 12 million tons, but the year after it dropped, and from 2005 to 2008 was on the levels between 9,5 and 11,5 million tons per year. In a five-ten years perspective, the total available capacity from Russian arctic oil export terminals can reach the level of 100 million tons/year (Bambulyak and Frantsen, 2009). Therefore, the risk of large accidents with oil tankers will increase in the years to come, unless considerable measures are imposed to reduce such risk.

Tourism is one of the largest and steadily growing economic sectors world-wide. Travels to the far north have increased considerable during the last 15 years, and there are currently nearly one million tourists annually.

The high biodiversity of the oceans represents a correspondingly rich source of chemical diversity, and there is a growing scientific and commercial interest in the biotechnology potential of Arctic biodiversity. Researchers from several nations are currently engaged in research that could be characterised as bio-prospecting.

Aquaculture is growing along the coasts of northern Norway and Russia, and there are several commercial fish farms producing salmonids (salmon, trout), white fish (mainly cod) and shellfish.

Ocean acidification is greater and happening faster than any previous acidification process experienced in millions of years. The absorption of $\mathrm{CO}_{2}$ generally goes faster in colder waters and thus will rapidly affect the Barents Sea.

### 1.2 State and expected situation of the ecosystem (Figures 1.2-1.10, Tables 1.3-1.6, 1.9)

### 1.2.1 Climate

## Atmospherical conditions

In 2009, the weather over the North Atlantic was determined by cyclonic activity throughout the year, and northerly and easterly winds prevailed over the Barents and northern Norwegian Seas. In winter, spring and autumn, air temperature averaged over the western and eastern parts of the Barents Sea was higher than normal, with maximum positive anomalies $\left(3.9-4.1^{\circ} \mathrm{C}\right)$ in the eastern Barents Sea in January and March. In summer, positive anomalies did not exceed $1^{\circ} \mathrm{C}$, and small negative anomalies were observed in some months

## Water temperature

In general the temperatures in the entire Barents Sea in 2009 was still high (about 0.5$1.0^{\circ} \mathrm{C}$ above the long-term average), and at about the same levels as in 2008. At the end of the year the temperature in the Atlantic water masses was increasing again. In the beginning of 2010 the temperature decreased again, but is still above the longterm mean.

Sea surface temperature (SST) in the Barents Sea showed much of the same variations as the air temperatures. In winter, due to the warmer-than-usual air masses over the central and eastern Barents Sea and therefore the less-than-usual atmospheric cooling, the SST was higher than normal, with maximum positive anomalies $\left(1.0^{\circ} \mathrm{C}\right)$ in the central part of the sea. In the western and north-western Barents Sea, on the contrary, the SST was lower than normal throughout most of the year, with maximum negative anomalies $\left(-0.5^{\circ} \mathrm{C}\right)$ in April and July. The weaker-than-usual spring-and-summer warming caused decreasing SST anomalies. From June to August, negative anomalies of SST were observed in most of the sea. In autumn, SST anomalies increased due to the intensification of cyclonic activity and warm air-masses transport; maximum positive anomalies of SST (up to $1.6^{\circ} \mathrm{C}$ ) were found in the southern areas in November.

Development in the coastal waters is measured at the Ingøy fixed station, and show that during 2009 the surface temperature was only slightly above normal through most of the year except in late fall/early winter 2009/2010. In the deeper waters (at 250 $\mathrm{m})$, which is strongly influenced by Atlantic Water, the temperature was above normal throughout the year. In both the surface and deeper layers, the temperature increased (relative to the normal) in late fall 2009/early winter 2010, but decreased again in spring 2010, with surface temperatures around and deeper layers still slightly above the long term mean.

The Fugløya-Bear Island and Vardø-North sections, which capture all the Atlantic Water entering the Barents Sea from south-west, showed temperatures close to $0.5^{\circ} \mathrm{C}$ above the long-term mean in early 2009 (Figure 1.2). This is lower than the last 5-6 winter, and is due to lower air temperatures causing more intense heat loss in combination with weak inflow of Atlantic Water. Over the year the temperatures increased, and in October 2009 the temperature in south-west was $0.9^{\circ} \mathrm{C}$ above the long-term
mean. The annual mean temperature in 2009 was close to the year of 2008. In the beginning the temperature at the Vardø-North decreased again to $\sim 0.5{ }^{\circ} \mathrm{C}$ above the long term mean.

Temperature in the upper 200 m layer in the southern Barents Sea (Kola section) was higher than normal throughout the year of 2009, and, during the second half of the year, it was higher than in 2008 (Figure 1.3). At the beginning of the year, the weaker-than-usual seasonal cooling caused an increase in positive temperature anomalies (by $0.1-0.3^{\circ} \mathrm{C}$ ) in the Atlantic water compared to December of 2008. The positive anomalies changed slightly during the first half of the year, then they decreased to September due to easterly and northeasterly winds prevailed in spring and summer. During autumn, temperature anomalies in the main warm currents increased again due to the intensification of cyclonic activity and air-mass transport from the west. By December, temperature anomalies exceeded $1.0^{\circ} \mathrm{C}$ in all parts of the Kola Section, and the highest December temperature for the period from 1951 to the present was observed in the Murman Current. The annual temperature in the Murman Current in 2009 was typical of anomalous warm years and close to that of 2008.

Temperature in the bottom layer of the Barents Sea in August-September 2009 was typical of warm and anomalous warm years. Positive temperature anomalies were observed in most of the surveyed area and were, on average, $0.3-1.0^{\circ} \mathrm{C}$. The largest positive temperature anomalies $\left(>1.5^{\circ} \mathrm{C}\right)$ were observed in the eastern Barents Sea, in the areas adjacent to the Eastern Basin (Figure 1.4). Compared to 2008, the volume of cold Arctic waters increased significantly in the northern Barents Sea, and for the first time in the last three years waters with negative temperature were found in the Eastern Basin. So, in comparison with the previous year, it caused decrease in the spatially averaged bottom temperature of the surveyed area except the southern Barents Sea occupied by the Murman Current and the Central branch of the North Cape Current. In the beginning of 2010 the bottom temperatures in the south and southwestern parts were higher than in the same period in 2009, while they were lower in the deep central parts.

According to computations with a prediction model, based on harmonic analysis of the Kola Section temperature time series, the temperature of the Atlantic water in the Murman Current in 2010-2011 is expected to decrease to values typical of warm years, namely to $4.5 \pm 0.5^{\circ} \mathrm{C}$ (with anomaly of $+0.6^{\circ} \mathrm{C}$ ) in 2010 and to $4.4 \pm 0.5^{\circ} \mathrm{C}$ (with anomaly of $+0.5^{\circ} \mathrm{C}$ ) in 2011. The years of 2010 and 2011 are similar to 1989, 1991, 2001 and 2002.

## Salinity

The salinity variations show a close resemblance to temperature, although not completely. In Fugløya-Bear Island the salinity has been decreasing since 2006, while in Vard $\varnothing-\mathrm{N}$ it has increased over the last years. Salinity in the Atlantic water masses in 2009 was still high compared to the long term trend.

## Inflow of Atlantic water

The volume flux of Atlantic Water flowing into the Barents Sea is predominantly barotropic, with large fluctuations in both current speed and lateral structure. In general, the current is wide and slow during summer and fast, with possibly several cores, during winter. The mean transport of Atlantic Water into the Barents Sea for the period 1997-2009 is $2 \mathrm{~Sv}\left(\mathrm{~Sv}=10^{6} \mathrm{~m}^{3} \mathrm{~s}^{-1}\right)$ with an average of 2.2 Sv during winter and 1.8 Sv during summer. During years in which the Barents Sea changes from cold
to warm marine climate, the seasonal cycle can be inverted. Moreover, an annual event of northerly wind causes a pronounced spring minimum inflow to the western Barents Sea; at times even an outward flow.

The time series of volume transport reveals fluxes with strong variability on time scales ranging from one to several months (Figure 1.5). The strongest fluctuations, especially in the inflow, occur in late winter and early spring, with both maximum and minimum in this period. The recirculation seems to be more stable at a value of something near 1 Sv , but with interruptions of high outflow episodes.

The volume flux varies with periods of several years, and was significantly lower during 1997-2002 than during 2003-2006. The year of 2006 was a special year as the volume flux both had a maximum (in winter 2006) and minimum (in fall 2006). Since then the inflow has been low, particularly during spring and summer. The inflow in 2009 was much as in 2007 and 2008; moderate during winter followed by a strong decrease in spring. In early summer 2009 the flux was close to 1.5 Sv below the average. As the observational series still only have data until summer 2009, it cannot give information about the situation in fall 2009 and early winter 2010. There is no significant trend in the observed volume flux from 1997 to summer 2010.

## Ice conditions

The variability in the ice coverage in the Barents Sea is linked to the temperature of the inflowing Atlantic water, the northerly winds, and import of ice from the Arctic Ocean and the Kara Sea. The ice has a response time on temperature changes in the Atlantic inflow (one-two years), but usually the sea ice distribution in the western Barents Sea respond a bit quicker than in the eastern part. Due to the high temperatures there has been little ice in the last years (Figure 1.6). During the period 20032006 the winter ice edge had a substantial retreat towards north-east, but since then the ice area has increased.

For the first eight months of the year of 2009, the sea ice extent in the Barents Sea was less than normal, but more than in 2008. In comparison with the previous year, the ice coverage (expressed as a percentage of the sea area) was $10-18 \%$ more in JanuaryMay, $5-9 \%$ more in June and August and the same in July. Ice melting in summertime was more intensive than in 2008. By July, the south-eastern Barents Sea was ice-free, which is almost one month earlier than in 2008. Ice formation started in the northernmost sea only at the end of October. In October, the ice coverage was $13 \%$ less than normal and $5 \%$ less than in 2008. By December, the ice coverage of the Barents Sea was still lower than normal but higher than in 2008, a situation that continued into the beginning of 2010.

It is expected that there will be slightly less or around average ice conditions in 2010.

## Hydrochemical conditions

According to the chemical observations along the Kola Section in 2009, some decrease in oxygen saturation of the bottom layer was found in the southern Barents Sea compared to 2008: the oxygen saturation anomaly averaged from January to October was $-0.24 \%$ in 2009 , and $0.78 \%$ in 2008. Negative anomalies prevailed at the beginning of the year, while small positive anomalies prevailed in summer and autumn.

### 1.2.2 Phytoplankton

In Norwegian waters there was not observed any large aberration in the annual succession in the phytoplankton along the fixed transect (Vardø - North and FugløyaBear Island) in 2009. The spring bloom occurred from mid March to mid April within the "normal" period of the spring bloom at the Bear Island transect. The bloom starts in the coastal waters "spreading" out into the open areas. In April the diatoms were dominating. During summer the phytoplankton was compound of small flagellates, dinoflagellates, and at some stations diatoms. During autumn larger dinoflagellates was common, however, at some stations diatoms had moderate to high abundance.

### 1.2.3 Zooplankton

The mesoplankton biomass measured in August-September 2009 was clearly below the long-term mean in the Norwegian sector but with slightly higher values along the border to the Russian zone. A particular feature in 2009 is the very high biomass found in the Russian sector north of $75^{\circ} \mathrm{N}$ and east of $40^{\circ} \mathrm{E}$. The average zooplankton biomass in the western and central Barents Sea in 2009 was 5.87 g dry weight m ${ }^{-2}$ compared to 6.48 g in 2008 and 7.13 g in 2007 (Figure 1.7).

The macroplankton survey conducted in autumn and winter 2009 showed that on average, abundance of euphausiids in the west and northwest of the sea was close to the level of 2008 (Figure 1.8). However, in the center, east and coast areas the abundance indices of krill increased 1.5-2 times compared to 2008. In total the macroplankton survey showed that the abundance indices of euphausiids were above than the long-term mean.

The average zooplankton abundance in 2009, together with the considerable decline observed since 2006, suggest that the condition for local production is less favourable for 2010. The total production will probably depend largely on the magnitude of zooplankton advection from the Norwegian Sea. The macroplankton feeding conditions for planktivorous fish in 2010 is expected to be similar to 2009.

The abundance of gelatinous zooplankton, caught by pelagic trawling, show a lower abundance of gelatinous zooplankton in 2009 compared to 2008. Both in 2008 and in 2009, the distribution of "jellyfish" also showed a considerable overlap with regions poor in mesozooplankton biomass.

### 1.2.4 Northern shrimp

According to the Russian-Norwegian ecosystem survey in August - September 2009 the largest catches of the northern shrimp were recorded in the eastern and northern Barents Sea and north of Spitsbergen. The investigations of 2009 showed that the total stock of the northern shrimp increased compared to last year.

### 1.2.5 Fish

The current and expected situation of the commercial stocks in the Barents Sea addressed by the AFWG is given in later chapters. Therefore focus in this subchapter is on other main species that interacts with the AFWG stocks, and on the role of the AFWG species in an ecosystem perspective (e.g. as predators). Special attention is given when there are deviations from the general situation. An overview of the development of pelagic and demersal stocks is given in Figures 1.9 and 1.10.

## NEA cod consumption

The food consumption of cod in 1984-2009, based on data from the Joint RussianNorwegian stomach content data base, is presented in Table 1.3-1.4. The main prey items in 2009 were capelin, polar cod, krill, haddock, herring, shrimp, cod and amphipods. In comparison with 2008 the importance of capelin and herring has increased while the importance of krill and shrimp has decreased. The consumption calculations made by IMR show that the total consumption by age 1 and older cod in 2009 was about 6 million tonnes (Table 1.3), while similar calculations by PINRO gave about 5 million tonnes. According to calculations by IMR and PINRO the consumption per cod was about the same in 2009 as in 2008 (Tables 1.5-1.6).

## Blue whiting and polar cod

Based on the most recent estimates of fishing mortality and SSB, ICES classifies the stock as having full reproductive capacity, and being harvested sustainably. SSB increased to a historical high in 2004 but has decreased since, and is expected to be just above $B_{p a}$ in 2011. The estimated fishing mortality is slightly below $\mathrm{F}_{\mathrm{pa}}$. Recruitment in 1995-2004 was at a much higher level than earlier, but the 2005 and later year classes seem to be poor. Total landings in 2008 were 1.3 mill. tonnes, which is lower than in 2007. Blue whiting is not fished in the Barents Sea.

The high abundance of blue whiting in the Barents Sea in 2004-2007 may be due to a large stock size in this period combined with high temperature. Blue whiting has been observed in the western and southern Barents Sea for many years, but never in such quantities, and never as far east and north in this area as in 2004-2007. In autumn 2009, the acoustic abundance of blue whiting was estimated to 0.3 million tonnes, which is higher than in 2008, but still low. Also, the swept area estimate of blue whiting in winter 2010 was the lowest in the time series, which go back to 2001. Thus, the abundance of blue whiting in the Barents Sea is expected to stay at a low level until the recruitment to the stock increases again.

The polar cod stock is presently at a high level. Norway took some catches of polar cod in the 1970s and Russia has fished on this stock more or less on a regular basis since 1970. The stock size has been measured acoustically since 1986 and the stock has fluctuated between 0.1-1.9 million t . In 2009, the stock size was measured to about 0.9 million $t$., which is below the estimate obtained in 2008. The natural mortality rate in this stock seems to be very high, and this is explained by the importance of polar cod as prey for cod and different stocks of seals.

## Herring and capelin

Based on the most recent estimates of SSB and fishing mortality, ICES classifies the stock as having full reproductive capacity and being harvested sustainably. The 1998, 1999, 2002 and 2004 year classes dominate the current spawning stock which is estimated to 12.2 million $t$ in 2010. Preliminary indications show that the year classes 2005-2009 are below average. Therefore the abundance of herring in the Barents Sea is believed to be at a relatively low level in 2010. This stock has shown a large dependency on the occasional appearance of very strong year classes. In recent years the stock has tended to produce strong year classes more regularly. However, as no strong year classes have been produced since 2004, the stock is expected to decline. Norwegian spring-spawning herring is fished along the Norwegian coast and in the Norwegian Sea, but not in the Barents Sea. However, juveniles from this stock play an important part role in the ecosystem in the Barents Sea.

The capelin stock size is at a level somewhat above average. Based on the most recent estimates of SSB and recruitment ICES classifies the stock as having full reproductive capacity. The maturing component in autumn 2009 was estimated to be 2.3 mill t ., and SSB 1st April 2010 was predicted to be at 0.52 mill t. The spawning stock in 2010 consisted of fish from the 2006 and 2007 year classes, but the 2006 year class dominated. The survey estimate at age 1 of the 2008 year class is somewhat below the long-term average. Observations during the international 0-group survey in AugustSeptember 2009 indicated that the 2009 year class is below average.

The estimated annual consumption of capelin by cod has varied between 0.2 and 3.0 million $t$ over the period 1984-2009. Young herring consume capelin larvae, and this predation pressure is thought to be one of the causes for the poor year classes of capelin in the periods 1984-1986, in 1992-1994, and from 2002-2005.

## Non-commercial species

Thorny skate (Amblyraja radiata) was quite widely distributed in the Barents Sea except for the south eastern and north eastern regions, as in 2008. The observed abundance of this species was higher than in 2008. The thorny skate preferred to keep in a wide range of depths from 50 down to 300 meters.

Northern skate (Amblyraja hyperborea) was distributed in the northeast part of the Barents Sea and along the shelf slope to the west of Spitsbergen. It was mainly found in the depth range 200 to 300 meters.

Plaice (Pleuronectes platessa) was distributed in a range of depths from 50 down to 100 meters) on northwest from Kanin peninsula.

According to observations in 2009 the tendency of expansion of Norway pout (Trisopterus esmarkii) in the Barents Sea is continuing. Main density concentrations of Norway pout were registered in the south-western areas. At the same time, along the warm Spitsbergen current, Norway pout was observed until $81^{\circ}$ N. Along coastal North Cape current Norway pout were distributed eastward up to $47^{\circ} \mathrm{E}$. It seems like Norway pout have occupied the blue whiting distribution area after this species declined.

In the ecosystem survey in 2009 there were both new species to the area and recordings of rare species in the area of observation. Some of these species have their main distribution in the warm waters of the Norwegian Sea (Molva molva, Schedophilus medusophagus) or in the cold waters of the Kara Sea (Arctogadus glacialis) bordering the Barents Sea.

### 1.2.6 Marine mammals

## Harp Seal

Since 1998 the abundance of harp seal pup production in the White Sea has been sharply reduced, according to the PINRO aerial surveys. However the decrease in the harp seal pup production abundance has become slower recently and even some slight increase has been observed. The abundance of harp seal pups in the whelping patches in 2009 calculated using the data from aerial surveys was more than two times lower, compared to the data obtained for 2000-2003.

One of the key factors, which caused the reduction in the harp seal pup abundance in 2004-2009, was the diminished ice extent due to warming. The changed ice conditions were responsible for the redistribution of animals in the pup period. Abnormal ice
conditions in the White Sea in 2005-2009 possibly also led to higher natural mortality of pups.

The decrease in the abundance of harp seal pup production leads to a reduction of the whole harp seal population (the model estimate for 2009-1.2 million animals).

## Predation by mammals

Analyses of consumptions by marine mammals in the Barents Sea for 2009 are not available. Last estimates are shown in Table 1.9.

### 1.2.7 Future long-term trends

This section is a short version of Stiansen et al (2009).
Air temperatures have increased almost twice as fast in the Arctic than the global average over the last 50 years. Models predict that air temperatures will continue to increase considerably. With the accelerated increase in air temperatures it is predicted that summer sea ice will disappear. Polar Front that separates the cold Arctic and warm Atlantic waters will move farther north and east. Although long-term climate projections are associated with considerable uncertainty, it is highly likely, however, that any significant warming will cause shifts in species ranges and changes in their production. The expected northward extension of warm Atlantic water will lead in general to that temperate zooplankton would shift northward while ice fauna, such as the large amphipods would diminish due to a massive loss of habitat because of the disappearance of multi-year ice (Skjoldal et al., 1987; Loeng et al., 2005). Ellingsen et al. (2008) also predicted that the Atlantic zooplankton production, primarily Calanus finmarchicus, would increase by about $20 \%$ and spread farther eastward while the Arctic zooplankton biomass would decrease significantly (by 50\%) resulting in an overall decrease in zooplankton production in the Barents Sea.

A number of fish species, e.g. cod and capelin, will likely have a more northern and/or eastern distribution and boreal species such as blue whiting and mackerel may become common in the Barents Sea. These changes will likely result in potentially large changes in community composition and it is possible that the structure of the ecosystem may shift irreversibly. In addition, sea ice extent will be reduced, and this will have a negative impact on ice-dependent flora and fauna, such as polar bears. Reduction in sea ice extent may also lead to increased primary productivity, if nutrient supply is not reduced significantly due to increased stratification in the water column. An increase in primary productivity coupled with other positive effects of increased temperature on fish growth and reproduction, may cause productivity of cod, haddock and other commercially important species to increase. However, negative effects on prey species may also occur. Thus, overall effects on fish productivity are hard to predict.

Similarly, the many complex ways in which species interact creates considerable uncertainty in any set of predictions as to what the overall response of climate warming to the ecosystem will be. If warming causes phytoplankton to increase, this is expected to result in an overall increase in fish production. For example, model studies show that higher primary production tends to lead to an increase in cod recruitment in the Barents Sea (Svendsen et al., 2007). Higher temperatures should also lead to improved growth rates of the fish and together with increased recruitment is expected to lead to increased fish yields (Drinkwater, 2005; Stenevik and Sundby, 2007). The results of long-term simulations by the STOCOBAR model show that a temperature increase of $1-4 C^{\circ}$ in the Barents Sea will lead not only to acceleration of cod
growth and maturation rates, but also to increase in cannibalism (Stiansen et al. 2009). Increased overall production is expected to produce increased catches of cod, haddock and other species (ACIA, 2005). Cod are expected to spawn farther north and new spawning sites will likely be established (Sundby and Nakken, 2008; Drinkwater 2005). With increasing temperatures, temperate benthic species are expected to become more frequent and the species composition of the benthos will change. Such changes will affect benthic production (i.e. food for demersal fishes and other vertebrates) and may therefore have considerable management implications. Polar bears, ringed seals, bearded seals, harp seals and hooded seals are all dependent on sea ice. It is the primary foraging habitat for polar bears, and a resting and breeding habitat for all of these seals. Additionally, some of the seals feed on ice-associated prey. As a result of climate warming and the associated loss of sea ice, distribution and abundance of these species are expected to decrease in the Barents Sea.

Along with climate change should mention that anthropogenic emissions of $\mathrm{CO}_{2}$ are causing acidification of the world oceans because $\mathrm{CO}_{2}$ reacts with seawater to form carbonic acid. Currently, acidity has increased by about $30 \%$ (reduction in pH by about 0,1 units). In 2100, pH reductions in the order of $0.2-0.3$ units are predicted. This will significantly reduce the ability of organisms to build calcium carbonate shells and skeletons and it might also have other effects on organisms. The direct effects are expected to be most pronounced for phytoplankton, zooplankton and benthos. Fish, seabirds and marine mammals can be affected indirectly, possibly making ocean acidification one of the most important anthropogenic drivers in the Barents Sea in the future.

### 1.3 Description ofo the Barents Sea fisheries and its effect on the ecosystem (Tables 1.10-1.11, Figures 1.11-1.16

Description of the Barents Sea fisheries and its effect on the ecosystem (Tables 1.101.11, Figures 1.11-1.16)

Fishing is the largest human impact to the fish stocks in the Barents Sea, and thereby the functioning of the whole ecosystem. However, the observed variation in both fish species and ecosystem is also impacted by other effects such as climate and predation. Open ocean fisheries in the Barents Sea started in the beginning of the $20^{\text {th }}$ century with the development of trawling technology. At present there is a multinational fishery operating in the Barents Sea using different fishing gears and targeting several species. The largest commercially exploited fish stocks (Northeast Arctic cod, haddock and saithe) are now harvested within sustainable limits and have full reproductive capacity. However, some of the smaller stocks (golden redfish, beaked redfish and coastal cod) are overfished, and damage to benthic organisms and habitats from trawling has been documented. Overcoming these problems and further developing our understanding of the effects of fisheries in an ecosystem context are important challenges for management.

### 1.4 General description of the fisheries

The major demersal stocks in the Northeast Arctic include cod, haddock, saithe, and shrimp. In addition, redfish, Greenland halibut, wolffish, and flatfishes (e.g. long rough dab, plaice) are common on the shelf and at the continental slope, and ling and tusk at the slope and in deeper waters. In 2008, catches of nearly 900 thousand tonnes are reported from the stocks of cod, haddock, saithe, redfish, and Greenland halibut, which is an increase of $10 \%$ as compared to the year before. An additional catch of
about 40000 tonnes was taken from the stocks of wolffish and shrimp. The annual fishing mortalities F (the mortality rate is linked to the proportion of the population being fished by $1-e^{-}$F for the assessed demersal fish stocks show large temporal variation within species and large differences across species from 0.1 ( $\approx 10 \%$ mortality) for some years for Sebastes marinus to above 1 ( $\approx 63 \%$ mortality) for some years for cod (Figure 1.11a). The current harvest rate relative to the maximum levels above which the fishing mortality over time may impair the recruitment is shown in Figure 1.11b. Of the analytically assessed demersal stocks in the Barents Sea it is currently only golden redfish (Sebastes marinus) which is harvested above this critical level.

The major pelagic stocks are capelin, herring, and polar cod. There was no fishery for capelin in the area in 2004-2008 due to the stock's poor condition, but in 2009 and 2010 the stock is again sufficiently sound to support a quota of 390000 and 360000 tonnes, respectively.

Russia, as the only nation currently fishing polar cod, fished 8190 tonnes polar cod in 2008. Norwegian spring spawning herring is the largest stock inhabiting the Northeast Arctic with its spawning stock estimated to 12.6 million tonnes in 2009. 1.5 million tonnes were fished from this stock in 2008, of which about 280000 tonnes were caught near the Norwegian coast in the south-western part of the Barents Sea. The highly migratory species blue whiting and mackerel extend their feeding migrations into this region, and in 2007 about 65000 tonnes mackerel and 120000 tonnes blue whiting were caught in the area, none of this, however, within the Barents Sea. Species with relatively small landings include salmon, Atlantic halibut, hake, pollack, whiting, Norway pout, anglerfish, lumpsucker, argentines, grenadiers, flatfishes, dogfishes, skates, crustaceans, and molluscs.

The most widespread gear used in the central Barents Sea is bottom trawl, but also long line and gillnets are used in the demersal fisheries. The pelagic fisheries use purse seine and pelagic trawl. Other gears more common along the coast include handline and Danish seine. Less frequently used gears are float line (used in a small but directed fishery for haddock along the coast of Finnmark, Norway) and various pots and traps for fish and crabs. The gears used vary with time, area and country, with Norway having the largest variety because of the coastal fishery. For Russia, the most common gear is bottom trawl, but a longline fishery mainly directed at cod and wolffish is also present. The other countries mainly use bottom trawl.

For most of the exploited stocks an agreed quota is decided (TAC), and also a number of additional regulations are applied. The regulations differ among gears and species and may be different from country to country, and a non-exhaustive list as well as a description of the major fisheries in the Barents Sea by species can be found in Table 1.10.

From 2011 onwards, the minimum mesh size for bottom trawl fisheries for cod and haddock will be 130 mm for the entire Barents Sea (at present the minimum mesh size is 135 mm in the Norwegian EEZ and 125 mm in the Russian EEZ). This change is not expected to have a significant impact on the total exploitation pattern for these stocks, thus a recent average exploitation pattern is used in the predictions.

### 1.4.1 Mixed fisheries

The demersal fisheries are highly mixed, usually with a clear target species dominating, and with low linkage to the pelagic fisheries (Table 1.11). Although the degree of mixing may be high, the effect of the fisheries varies among the species. More specifically, the coastal cod stock and the two redfish stocks are presently at very low levels.

Therefore, the effect of the mixed fishery will be largest for these stocks. In order to rebuild these stocks, further restrictions in the regulations should be considered (e.g. closures, moratorium, and restrictions in gears).
Successful management of an ecosystem includes being able to predict the effect of a mixed fishery on the individual stocks, and ICES is requested to provide advice which is consistent across stocks for mixed fisheries. Work on incorporating mixed fishery effects in ICES advice is ongoing and various approaches have been evaluated (ICES 2006/ACFM:14). At present such approaches are largely missing due to a need for improving methodology combined with lack of necessary data. However, technical interactions between the fisheries can be explored by the correlation in fishing mortalities among species (Figure 1.12). The correlation in fishing mortality is positive for Northeast Arctic cod and coastal cod, and for haddock and coastal cod confirming the linkage in these fisheries. There is also a significant relationship between saithe and Greenland halibut although the linkage in these fisheries is believed to be low (Table 1.11). The relationships between the other fishing mortalities are scattered and inconclusive. In case of strong dependencies in fishing mortalities this method can, in principle, be used to produce consistent advice across species concerning fishing mortality. It is however too simple since this correlation is influenced by too many confounding factors whose effect cannot be removed without a detailed analysis of data with a higher resolution (e.g. saithe and Greenland halibut, and changes in stock distribution (ICES 2006/ACFM:14).

A further quantification of the degree of mixing and impact on individual stocks requires detailed information about the target species and mix per catch/landing and gear. Such data exist for some fleets (e.g. the trawler fleet), but is incomplete for other fleets. The Russian and Norwegian trawl fleet catches show spatial and temporal differences in both composition and size as well as large differences between countries (Figures 1.13-1.16). In the north eastern part of the Barents Sea the major part of the Russian catches consists of cod, whereas the Norwegian catches include a large proportion of other species (mainly shrimp). In the most western part of the Barents Sea, the Norwegian catches consist of Sebastes mentella and Greenland halibut in addition to cod, whereas the Russian catches mainly consist of cod and haddock. The main reason for this disparity is the difference in spatial resolution of the data; the Norwegian strata system extends further west and thus covers the fishing grounds of Greenland halibut, whereas the Russian strata does not. The Norwegian trawl fishery along the Norwegian coast includes areas closer to the coast and is also more southerly distributed where other species are more dominant in the catches (e.g. saithe).

Estimates of unreported catches of cod and haddock in 2002-2008 indicate that this has been a considerable problem which now seems to be decreasing. According to the report from the Norwegian-Russian analytical group the total catches of both cod and haddock reported to AFWG are very close (within $1 \%$ ) to the estimates made by this group. Thus it was decided to set the IUU catches for 2009 to zero (see chapter 0.4). A continuous control and surveillance of this problem is necessary. Discarding of cod and haddock (and in some years also saithe) is thought to be significant in periods, although discarding of these, and a number of other species, is illegal in Norway and Russia. Data on discards are scarce, but attempts to obtain better quantification are ongoing.

## Fleet composition (groundfish and pelagic species)

Figure 1.17 shows the main fleets catching bottom and pelagic fishes in the Barents Sea and Svalbard (Spitsbergen archipelago) areas. The pelagic fishery is only conducted by Russia and Norway where both countries target the capelin. Russia has, in addition, fished polar cod with pelagic trawl (Norway has not fished this species since the early 1980s), and Norway has in recent years fished some legal sized herring in a restricted coastal purse seine fishery inside 4 nautical miles off Finnmark. Further in the south western part of the Barents Sea (south-west of a line between Sørøya and Bear Island), extending into the Norwegian Sea, an international herring fishery has been open in some seasons.

The Norwegian groundfish fishery is much more diverse compared to Russia and other countries regarding the number of fleets. The trawler fleet itself is also rather diverse both within and between countries. In the Norwegian groundfish fishery several other gears are also used in addition to trawl. The gear composition also depends on which groundfish species the fishery targets. The Norwegian bottom trawl fleet catch about $30 \%$ of the Norwegian cod catch, about $40 \%$ of the haddock, and more than $40 \%$ of the Norwegian saithe and Greenland halibut catches. The Russian bottom trawl fleet catch about $100 \%$ of the Russian saithe catch, about $95 \%$ of cod and haddock, $90 \%$ of the Russian Greenland halibut catch and about $37 \%$ of wolffishes. Other countries fishing groundfish in these waters only use trawl, incl. some pairtrawling. It is mandatory in all groundfish trawl fisheries to use sorting grid to avoid catching undersized fish. The one and only exception from this rule is within an area in the southwestern part of the Barents Sea during 1 January - 30 April where trawling without sorting grids is permitted to catch haddock.

## Impact of fisheries on the ecosystem

In order to conclude on the total impact of trawling, an extensive mapping of fishing effort and bottom habitat would be necessary. In general, the response of benthic organisms to disturbance differs with substrate, depth, gear, and type of organism (Collie et al. 2000). Seabed characteristics from the Barents Sea are only scarcely known (Klages et al. 2004) and the lack of high-resolution ( $\pm 100 \mathrm{~m}$ ) maps of benthic habitats and biota is currently the most serious impediment to effective protection of vulnerable habitats from fishing activities (Hall 1999). An assessment of fishing intensity on fine spatial scales is critically important in evaluating the overall impact of fishing gear on different habitats and may be achieved, for example, by satellite tracking of fishing vessels (Jennings et al. 2000). The challenge for management is to determine levels of fishing that are sustainable and not degradable for benthic habitats in the long run.

Fisheries in the Barents Sea do not only influence the targeted stocks. Due to strong species interactions fisheries removal of one stock may influence the abundance of other stocks. For example, herring collapses have positively influenced capelin abundance. Reduced stock sizes due to fisheries removal may also lead to changing migration patterns. Due to density dependent migrations, fish stocks cover greater areas and migrate longer distances when abundances are high compared to low. Fisheries also reduce the average fish size, age and age at maturity. The reduced size and age of the cod stock may actually have altered the ecological role of cod as top predators in the Barents Sea.

The qualitative effects of trawling have been studied to some degree. The most serious effects of otter trawling have been demonstrated for hard-bottom habitats domi-
nated by large sessile fauna, where erected organisms such as sponges, anthozoans and corals have been shown to decrease considerably in abundance in the pass of the ground gear. Barents Sea hard bottom substrata, with associated attached large epifauna should therefore be identified.

Effects on soft bottom have been less studied, and consequently there are large uncertainties associated with what any effects of fisheries on these habitats might be. Studies on impacts of shrimp trawling on clay-silt bottoms have not demonstrated clear and consistent effects, but potential changes may be masked by the more pronounced temporal variability in these habitats (Løkkeborg 2005). The impacts of experimental trawling have been studied on a high seas fishing ground in the Barents Sea (Kutti et al. 2005.) Trawling seems to affect the benthic assemblage mainly through resuspension of surface sediment and through relocation of shallow burrowing infaunal species to the surface of the seafloor.

Work is currently going on in the Arctic, jointly between Norway and Russia, exploring the possibility of using pelagic trawls when targeting demersal fish. The purpose is to avoid impact on bottom fauna and to reduce the mixture of other species. It will be mandatory to use sorting grids to avoid catches of undersized fish.

Lost gears such as gillnets may continue to fish for a long time (ghost fishing). The catch efficiency of lost gillnets has been examined for some species and areas (e.g. Humborstad et al. 2003; Misund et al. 2006; Large et al. 2009), but at present no estimate of the total effect is available. Ghost fishing in depths shallower than 200 m is usually not a significant problem because lost, discarded, and abandoned nets have a limited fishing life owing to their high rate of biofouling and, in some areas, their tangling by tidal scouring. Investigations made by the Norwegian Institute of Marine Research of Bergen in 1999 and 2000 showed that the amount of gillnets lost increases with depth and out of all the Norwegian gillnet fisheries, the Greenland halibut fishery is the metier where most nets are lost. The effect of ghost fishing in deeper water, e.g. for Greenland halibut, may be greater since such nets may continue to "fish" for periods of at least 2-3 years, and perhaps even longer (D. M. Furevik and J. E. Fosseidengen, unpublished data), largely as a result of lesser rates of biofouling and tidal scouring in deep water. The Norwegian Directorate of Fisheries has organised retrieval surveys annually since 1980. All together 10784 gill nets of 30 metres standard length (approximately 320 km ) have been removed from Norwegian fishing grounds during the period from 1983 to 2003.

Other types of fishery-induced mortality include burst net, and mortality caused by contact with active fishing gear, such as escape mortality (Suuronen 2005; Broadhurst et al. 2006; Ingólfsson et al. 2007). Some small-scale effects are demonstrated, but the population effect is not known.

The harbour porpoise is common in the Barents Sea region south of the polar front and is most abundant in coastal waters. The harbour porpoise is subject to by-catches in gillnet fisheries (Bjørge and Kovacs 2005). In 2004 Norway initiated a monitoring program on by-catches of marine mammals in fisheries.

Fisheries impact seabird populations in two different ways: 1) Directly through bycatch of seabirds in fishing equipment and 2) Indirectly through competition with fisheries for the same food sources.

Documentation of the scale of by-catch of seabirds in the Barents Sea is fragmentary. Special incidents like the by-catch of large numbers of guillemots during spring cod fisheries in Norwegian areas have been documented (Strann et al. 1991). Gillnet fish-
ing affects primarily coastal and pelagic diving seabirds, while the surface-feeding species will be most affected by long-line fishing (Furness 2003). The population impact of direct mortality through by-catch will vary with the time of year, the status of the affected population, and the sex and age structure of the birds killed. Even a numerically low by-catch may be a threat to red-listed species such as Common guillemot, White-billed diver and Steller's eider.

Several bird scaring devices has been tested for long-lining, and a simple one, the bird-scaring line (Løkkeborg 2003), not only reduces significantly bird by-catch, but also increases fish catch, as bait loss is reduced. This way there is an economic incentive for the fishermen to use it, and where bird by-catch is a problem, the bird-scaring line is used without any forced regulation.

In 2009, the Norwegian Institute for Nature Research (NINA) and the Institute of Marine Research in Norway started a cooperation to develop methods for estimation of bird by-catch. Preliminary reports from observers at sea trained by the institutes show that most of the fisheries have a minor impact on bird mortality.

### 1.5 Management improvement issues (Tables 1.12-1.15)

### 1.5.1 Overview

The availability of necessary ecosystem information is only one of the needed items for implementation of an ecosystem approach to management. Another needed element is the development of appropriate methods and instruments for incorporation of ecosystem information into stock assessment and harvest control rules.

This section summarizes ecosystem information that has the potential of being implemented in, and therefore improves, the advice for sustainable fishery management.

Management of fisheries is always based on decision-making under levels of uncertainty. Incorporating data on ocean climate, lower trophic level bio-production, as well as species interactions on higher trophic levels in catch recommendations for target species, should reduce the uncertainty of scientific recommendations for sustainable harvest levels.

### 1.5.2 Multispecies models

Development of multispecies models designed to improve fisheries management in the Barents Sea based on species interactions started in the mid-1980s. The first models developed were MULTSPEC, AGGMULT and SYSTMOD in IMR and MSVPA in PINRO (Tjelmeland and Bogstad, 1998; Hamre and Hatlebakk, 1998, Korzhev and Dolgov, 1999). In total, these models contained the species cod, capelin, herring, haddock, polar cod, shrimp, harp seal and minke whale. Even though further development of these models has been discontinued, they serve as predecessors to newly developed models, such as EcoCod, Bifrost, Gadget and STOCOBAR. Benefits of multispecies models include: improved estimates of natural mortality and recruitment; better understanding of stock-recruit relationships and variability in growth rates; alternatives views on biological reference points. Brief descriptions of the multispecies models are given below.

## EcoCod

The development of this model started in 2005 as the main task in the first stage of the joint PINRO-IMR Programme on Estimation of Maximum Long-Term Yield of North-East Arctic cod, taking into account the effect of ecosystem factors. This 10year research programme was initiated following a request from the RussianNorwegian Fishery Commission (Filin and Tjelmeland, 2005). EcoCod is a stepwise extension of a single species model for cod (CodSim; Kovalev and Bogstad, 2005), where cod growth, maturation, cannibalism and recruitment is modeled in a multispecies setting. Preliminary sub-models for cod growth, fecundity and malformation of eggs have been implemented in EcoCod.

## Bifrost

Bifrost (Boreal integrated fish resource optimization and simulation tool) is a multispecies model for the Barents Sea (Tjelmeland and Lindstrøm, 2005) with main emphasis on the cod-capelin dynamics. The prey items for cod are younger cod, capelin and other food. The predation model is estimated by comparing simulated consumption to that calculated from individual stomach content data using the dos Santos evacuation rate model with a parameterization where the initial meal size is excluded. The capelin availability partly shields the cod juveniles from cannibalism, and by including this effect, the recruitment relation for cod is significantly improved.
In prognostic mode, Bifrost is coupled to the assessment model for herring - SeaStar (Tjelmeland and Lindstrøm, 2005) - and the negative effect of herring juveniles on capelin recruitment is modeled through the recruitment function for capelin. Bifrost is also used to evaluate cod-capelin-herring multispecies harvest control rules.

## STOCOBAR

The STOCOBAR describes stock dynamics of cod in the Barents Sea, taking into account trophic interactions and environmental influence (Filin, 2007). It is designed as a tool for prediction and exploration of cod stock development as well as for evaluation of harvest strategies and recovery plans under different ecosystem scenarios. The STOCOBAR is an age-structured, a single-area and a single-fleet model with one year time steps. It includes a cod as predator on up to eight prey items: capelin, shrimp, polar cod, herring, krill, haddock, own young and other food. Species structure of the model is not permanent and it can be reduced from seven-species version to a simple version, which includes cod and capelin only. Recruitment function is used for cod only. Impact assessment of ecosystem factors on cod stock dynamics are based on «what if» scenarios. STOCOBAR is able to take uncertainties in future scenarios of temperature and capelin stock dynamics, in abundance and individual weight of cod at age 1 and in its fishing mortality rate into account. The first version of STOCOBAR was created at PINRO in 2001 and development of this model is continuing. The work on the development of the STOCOBAR model is part of the Barents Sea Case Study within the EU project UNCOVER (2006-2010) and the joint PINRO-IMR project (2004-2013) Optimal long-term harvest in the Barents Sea.

## GADGET

A multi-species Gadget age-length structured model ( www.hafro.is/gadget ; Begley and Howell, 2004, developed during the EU project dst² (2000-2003)), is being used for modeling the interactions between cod, herring, capelin and minke whale in the

Barents Sea as part of the EU projects BECAUSE, UNCOVER, DEFINEIT and FACTS. This is a multi-area, multi-species model, focusing on predation interactions within the Barents Sea. The predator species are minke whale, cod and herring, with capelin, immature cod, and juvenile herring as prey species. Krill is included as an exogenous food for minke whales (Lindstrøm et al. 2009). The cod model employed is based on the model presented at AFWG each year.

The modeling approach taken has many similarities to the MULTSPEC approach (Bogstad et al., 1997). Work is ongoing to enhance the modeling of recruitment processes during the EU projects FACTS and DEFINEIT. An FLR routine has been written that can run Gadget models as FLR Operating Models. This also gives the possibility of using Gadget as an operating model to test the performance of various assessment programs under a range of scenarios (Howell and Bogstad, 2010). In addition the Gadget multi-species model is being developed to assess the likely impact on medium-term population dynamics of oil-spill induced larval mortalities.

### 1.5.3 Statistical models

## Recruitment of commercial fish

Prediction of recruitment in fish stocks is essential for harvest prognosis stocks, both in a single-species and multi-species context. Traditionally, prediction methods have been based on spawning stock biomass and survey indices of juvenile fish and have not included effects of climate variability. Multiple linear regression models can be used to incorporate both climate and parental fish stock parameters. In order for such models to give predictions there need to be a a time lag between the predictor and response variables.

## Maturation of cod

The decrease in capelin stock biomass potentially impacts the maturation dynamics of Northeast Arctic cod by delaying the onset of maturation and/or increasing the incidence of skipped spawning. The relationship between weight- and length-at age shows that for a given length, weight-at-length is positively correlated with proportion mature-at-length for the period 1985-2001 (Marshall et al., 2004).

Estimates of weight-at-length were multiplied by the Russian liver condition index at length (Yaragina and Marshall, 2000) to derive estimates of liver weights in grams for cod at a standard length (see Marshall et al. 2004 for details of the calculation). This analysis indicated that for the period 1985-2001 there is a consistently significant, positive relationship between liver weight and proportion mature.

## Condition of fish

Relative body condition (the quantity of stored energy) is an important tool in understanding demographic variation and the ability of a population to respond to environmental stressors, varying food availability and competition. A high-resolution database was used to examine causes of variation in the condition of North-east arctic cod for the period 1967-2004, over annual and monthly timescales. Temperature was shown to have a positive impact on condition at both inter- and intra-annual timescales. Interannually, temperature may affect stock distribution, in particular its overlap with the capelin stock. At shorter timescales it is likely that temperature directly affects the metabolism of the cod. Intra-annually, the quantity of capelin in cod stomachs positively affected cod condition in the current and the preceding month for
all lengths of cod. This indicated a time lag between a change in food consumption and a subsequent change in condition, or 'latency'.

Results presented by Sandeman et al. (2008) point to the importance of the impact of varying temperature on condition. The effects of climate are likely to be particularly important where the species is close to its outer distribution area or where the animal is an ectotherm.

## Growth of fish

Large interannual variations in growth rate are observed for all commercial fish species in the Barents Sea. The most important causes are temperature change, density dependence and changes in prey availability. Variation in growth rate can contribute substantially to variability in stock biomass and can have a large impact on reproductive output. Regressions of weight at age of cod on temperature, capelin and the cod stock itself are used in EcoCod model.

Growth of the youngest capelin is correlated with abundance of the smallest zooplankton, whereas growth of older capelin is more closely correlated with abundance of the larger zooplankton. The developed regression equations have low determination coefficient, and are therefore not presented here. However, they may prove useful in the future when further developed.

## Reproductive potential

Morgan et al. (2009) explore the impact of four alternative indices of reproductive potential (RP) on perceptions of population productivity for eight fish populations across the North Atlantic. The four indices of RP included increasing biological complexity, adding variation in maturation, sex ratio, and fecundity. Perceptions of stock productivity were greatly affected by the choice of index of RP. Population status relative to reference points, RP per recruit, and projections of population size all varied when alternative indices of RP were used. There was no consistency in which index of RP gave the highest or lowest estimate of population productivity, but rather, this varied depending on how much variation there was in the reproductive biology of the population and the age composition. Estimates of sustainable harvest levels and recovery time for depleted populations can vary greatly depending on the index of RP.

### 1.5.4 Other models

## Consumption models

When calculating the prey consumption by a given predator, both the overall consumption level and the prey composition in the diet are used. The prey composition is usually derived from stomach content data, while the overall consumption level can be calculated using two approaches:

- A bioenergetic approach (as is usually the case for marine mammals and seabirds as predators)
- By combining data on stomach content weight with models for stomach evacuation rate, based on experiments.


## Ecosystem models

Ecosystem models may be useful for looking at how change in one species or ecosystem component is affecting whole or other parts an ecosystem, thereby identifying the most important inter-species/ functional group links and sensitivity of the ecosystem to changes to those. They are also useful for scenario testing (change in fishery pressure, climate change, and sudden pollution events. Special interesting are those models that have spatial resolution, like ATLANTIS and ECOPATH/ECOSIM.

Atlantis (Fulton et al., 2004a) is an ecosystem 3D box-model intended for use in management strategy evaluation (as described in de la Mare 1996, Cochrane et al. 1998, Butterworth and Punt 1999, Sainsbury et al. 2000). The overall structure of Atlantis is based around having multiple alternative submodels to represent each step in the management strategy and adaptive management cycles. It has been applied to multiple marine systems (from single bays to millions of square kilometers) in Australia and the United States. In autumn 2010 it will be implemented at IMR, and cover the area of the Barents Sea and the Norwegian Sea.

Another model that may have some utility for the future would be ECOPATH/ECOSIM. This model can use ecosystem survey data and expected biomass conversion rates to model systems. As a mass-balance model it can detect if there may be overlooked components to the ecosystem. The ECOPATH model system is used in many systems around the world. Versions of it have also been applied to the Barents Sea ecosystem (Blanchard et al. 2002, Dommasnes et al. 2002), though they are not run on an operational level.

### 1.5.5 Expected impact of ecosystem factors on stock dynamics

## Evaluation of natural potential of cod stock biomass changes based on temperature and capelin data

STOCOBAR long-term simulations show that impact of capelin on cod stock dynamics is dependent on temperature and cod stock state (WD21). Using these simulations the natural potential for changes in cod stock size may be identified based on temperature conditions and the state of cod and capelin stocks in the Barents Sea. A table for evaluating the level of natural potential for annual changes in fishable cod stock biomass was produced based on the simulated data (Table 1.12).

According to Table 1.12 and available data on temperature, cod and capelin stocks the potential for annual changes of cod fishable stock biomass in 2009 was low. The same situations will be in 2010 and 2011 based on expected temperature and capelin stock size. The resistance of cod stock to fishing pressure under these conditions will be medium and this does not imply high contributions to cod stock dynamics from capelin and temperature.

## Prediction of NEA cod recruitment.

Several statistical models, which use multiple linear regressions, have been developed for recruitment of North East Arctic cod. All models try to predict recruitment at age 3 (at 1 January), as calculated from the VPA, with cannibalism included. This quantity is denoted as R3. A collection of the most relevant models for AFWG is described below.

Stiansen et al. (2005) developed a model (JES1) with 2 year prediction possibility:

```
JES1: R3~ Temp(-3) + Age1(-2) + MatBio(-2)
JES2: R3~ Temp(-3) + Age2(-1) + MatBio(-2)
JES3: R3~ Temp(-3) + Age3(0) + MatBio(-2)
```

Temp is the Kola yearly temperature $(0-200 \mathrm{~m})$, Age1 is the winter survey bottom trawl index for cod age 1, and MatBio the maturing biomass of capelin. The number in parenthesis is the time lag in years. Two other similar models (JES2, JES3) can be made by substituting the term Age1(-2) with Age2(-1) and Age3(0), respectively (winter survey bottom trawl index for cod age 2 and age 3, respectively), This gives 1 and 0 year predictions, respectively.

Svendsen et al. (2007) used a model (SV) based only data from the ROMS numerical hydro-dynamical model, with 3 year prognosis possibility:
SV: R3~ Phyto(-3) + Inflow(-3)

Where Phyto is the modelled phytoplankton production in the whole Barents Sea and Inflow is the modelled inflow through the western entrance to the Barents Sea in the autumn. The number in parenthesis is the time lag in years. The model have not been updated since 2007.

The recruitment model (TB) suggested by T. Bulgakova (AFWG 2005 WD14, WD9) is a modification of Ricker's model for stock-recruitment defined by:

$$
\text { TB: R3~ m(-3) exp[-SSB }(-3)+N(-3)]
$$

Where R3 is the number of age3 recruits for NEA cod, $m$ is an index of population fecundity, SSB is the spawning stock biomass and $N$ is equal to the numbers of months with positive temperature anomalies (TA) on the Kola Section in the birth year for the year class. The number in parenthesis is the time lag in years. For the years before 1998 TA was calculated relatively to monthly average for the period 1951-2000. For intervals after 1998, the TA was calculated with relatively linear trend in the temperature for the period 1998-present. The model was run using two time intervals (using cod year classes 1984-2000 and year classes 1984-2004) for estimating the model coefficients. The models were not updated this year.

Titov (WD 22) and Titov et al. (WD 16 AFWG 2005) developed models with 1 to 4 year prediction possibility (TITOV1, TITOV2, TITOV3, TITOV4, respectively), based on the oxygen saturation at bottom layers of the Kola section stations 3-7 (OxSat), air temperature at the Murmansk station (Ta), water temperature: 3-7 stations of the Kola section (layer 0-200m) (Tw), ice coverage in the Barents Sea (I), spawning stock biomass (SSB), and the acoustic abundance of cod at age 1 and 2 , derived from the joint winter Barents Sea acoustic survey:

```
TITOV0: R3 \({ }^{1} \sim\) DOxSat \({ }^{2}(\mathrm{t}-13)+\) DOxSat(t-13) + ITa(t-39) \(+\operatorname{CodA3}(\mathrm{t}+1)+\mathrm{Tw}(\mathrm{t}-17)\)
TITOV1: R3¹ ~DOxSat²(t-13)+ DOxSat(t-13) + ITa(t-39) +CodA2(t-11) + Tw(t-17)
TITOV2: R3 \({ }^{2} \sim\) DOxSat²(t-13) - DOxSat(t-13) \(+\mathrm{ITa}(\mathrm{t}-39)+\operatorname{CodA1}(\mathrm{t}-23)+\mathrm{Tw}(\mathrm{t}-17)\)
TITOV3: \(\mathrm{RB}^{3} \sim \mathrm{OxSat}^{2}(\mathrm{t}-44)+\mathrm{ITa}(\mathrm{t}-39)+\lg \operatorname{CodC0}(\mathrm{t}-28)\)
TITOV4: R3 \({ }^{4} \sim\) OxSat \(^{2}(\mathrm{t}-44)+\mathrm{ITa}(\mathrm{t}-39)+\mathrm{SSB}(\mathrm{t}-36)\)
```

Where DOxSat(t-13) ~ Exp(OxSat(t-13)) - OxSat(t-38), ITa(t-39) ~I(t-39) +Ta(t-44). The number in parenthesis is the time lag in months, relative to 1 January at age 3 . The ITa index coincides in time with the increase of horizontal gradients of water temperatures in the area of the Polar Front (Titov, 2001). Some changes were brought in 2009 (AFWG 2009 WD 12). New equation (TITOV0) was added, 0-group abundance indic-
es, corrected for capture efficiency (CodC0) was entered instead of former indices in TITOV3.

Hjermann et al. (2007) developed a model with a one year prognosis, which have been modified by Dingsør et al (WD 19) to four models with 1-2 year projection possibility.

```
H1: \(\log (\mathrm{R} 3) \sim \operatorname{Temp}(-3)+\log (\) Age0 \()(-3)+\) BM \(_{\text {cod3-6 }} / \mathrm{ABM}_{\text {capelin }}(-2,-1)\)
H2: \(\log (\) R3 \() \sim\) Temp(-2) \(+\mathrm{I}(\) surv \()+\) Age1(-2) + BM \(_{\text {cod3-6 }} /\) ABM \(_{\text {capelin }}(-2,-1)\)
H3: \(\log (\) R3 \() \sim\) Temp(-1) + Age2(-1) + BM \(_{\text {cod3-6 }} /\) ABM \(_{\text {capelin }}(-1)\)
H4: \(\log (\) R3 \() \sim\) Temp(-1) + Age3(0)
```

Temp is the Kola yearly temperature $(0-200 \mathrm{~m})$, Age 0 is the 0 -group index of cod, Age1, Age2 and Age3 are the winter survey bottom trawl index for cod age 1,2 and 3, respectively, $\mathrm{BM}_{\text {cod } 3-6}$ is the biomass of cod between age 3 and 6 , and $A B M$ is the maturing biomass of capelin. The number in parenthesis is the time lag in years.

At AFWG 2008, Subbey et al. presented a comparative study (AFWG 2008 WD27) on the ability of some of the above models in predicting stock recruitment for NEA cod (Age 3). At the assessment meeting this year a WD by Dingsør et al. (WD 19) was presented, which investigated the performance of some of the mentioned recruitment models. Even though this work was well received by the working group it was decided not to change the procedure this year. However, it was strongly recommended that a Study Group should be appointed to look at criteria's for choosing/rejecting recruitment models suitable for use in stock assessment (see also chap 0.11).

The 2008 assessment agreed on using a combination of the best performing models according to Subbey at (AFWG 2008 WD27) for the age 3 predictions, names the "hybrid" model. One-year-ahead prognoses was given by the hybrids (Titov1, Titov3 and JES1), two-year-ahead (Titov2, Titov 3 and JES1) and three-year-ahead (Titov3) for the number of age 3 cod. Following the recommendation of the review group in 2008 this procedure was also conducted in the 2009 assessment.

At the 2010 assessment the model JES 1 was removed from the hybrid for the 2010 estimate only, due to a low age 1 index and thereby the model being out of its valid range for that prognosis year. Otherwise the hybrid model approach was similar to last year.

Table 1.13 show the estimates of all the available models, along with last year estimates.

## Cannibalism mortality for cod

Currently AFWG estimates of cod natural mortality caused by cannibalism based on data of the cod proportion in the cod diet is shown in Table 1.14. These data are used for estimation of cod consumed by cod and further for estimation of its natural mortality within the XSA (see section 3.4.2). Averaged natural mortality for last 3 years is used as predicted $M$ for next 4 years (section 3.7.1).

An alternative approach for prediction of NEA cod cannibalism was proposed by Kovalev (2004), based on the linear relationship between the natural mortality of cod at ages 3-5 and the biomass of cod spawning stock with minus 3-year lag. Using this approach the predicted natural mortality coefficient for cod, including cannibalism for recent years, seems to be higher compared to "the standard" assessment and prediction (Table 1.15). Because the mechanisms of cod SSB influence on the level of own young natural mortality on age 3-4 years is unclear, and because of this relation-
ship seems not to be in correspondence with observations over the last few years, the assessment group decided that this approach should not to be used for prediction before it will be further tested. Values for the years 2009 to 2012, predicted by the regression, are given in the Table 1.15.

### 1.5.6 Fishery induced evolution

There is a vital need for the fisheries science community to maintain sustainable fisheries ensuring the effective conservation, management and development of living aquatic resources. The precautionary approach was proclaimed and applied within the ICES community to meet (promote) these aims. This approach takes into account uncertainties relating to the size and productivity of the stocks. Uncertainties relating to fisheries induced evolution are most likely taken into consideration in case of a proper implementation of precautionary approach into responsible fishery.

The Study Group on Fisheries Induced Adaptive Change (SGFIAC) proposed to create evolutionary impact assessment (EvoIA), quantifying the evolutionary effects of management measures (ICES 2008/RMC:01; ICES 2009/RMC:03). It is a very complicated but promising task given that commercial fishery could act as a selective factor resulting in evolutionary response of exploited populations.

The papers published by the SGFIAC Group members concern basically probabilistic maturation reaction norms (PMRNs) estimations for different commercial stocks/species, and shift in cohort-specific PMRNs interpreted as a genetic change at the population level. It is rather difficult to test that findings directly as the genes associated with maturation have a polygenic nature. The strength and weakness of the PMRNs approach were discussed in detail in Theme Session issue of the Marine ecology progress series, 2007, vol. 335, 249- 310.

North east arctic cod stock demonstrates long-term trends in maturation as well as in demography of the stock and weight at length of fish. The historical trends could be caused both by genetic and plastic effects on maturation. Population density factors and environmental conditions can contribute to feeding success resulting in changing maturation rates in NEA cod for the time period investigated (Marshall and McAdam, 2007; Kovalev and Yaragina, 2009). The causes in a discontinuity of the decreasing trend observed in length for $50 \%$ maturation probability in the beginning of the 80's are unknown, but they are most likely non-genetic given that they occurred synchronously across age-classes (Marshall and McAdam, 2007).

More research is needed to evaluate underlined mechanisms of population changes including biological, physiological, ecological studies, not to mention genetic ones.

It takes a lot of time and efforts for the ICES community to implement the precautionary approach into a scientific/management practice. It is likely to take some time before the SGFIAC can evaluate and present some results applicable to test on real management measures recommendations. AFWG considers it premature at present to discuss any proposals of management measures (or reference points for fisheries management) in terms of fisheries induced evolution. Dialogues with scientists of the mentioned WG could also be carried out through the ICES Sharepoint.

### 1.6 Monitoring of the ecosystem (Figure 1.18, Tables 1.16-1.17)

Monitoring of the Barents Sea started already in 1900 (initiated by Nicolai Knipovich), with regular measurement of temperature in the Kola section. In the last 50 years regular observations of ecosystem components in the Barents Sea have been con-
ducted both at sections and by area covering surveys from ship and airplanes. In addition, there are conducted many long and short time special investigations, designed to study specific processes or knowledge gaps. Also, the quality of large hydrodynamical numeric models is now at a level where they are useful for filling observation gaps in time and space for some parameters. Satellite data and hindcast global reanalysed datasets are also useful information sources.

### 1.6.1 Standard sections and fixed stations

Some of the longest ocean time series in the world are along standard sections (Figure 1.18) in the Barents Sea. The monitoring of basic oceanographic variables for most of the sections goes back 30-50 years, with the longest time series stretching over one century. In the last decades also zooplankton is sampled at some of these sections. An overview of length, observation frequency and present measured variables for the standard sections in the Barents Sea is given in Table 1.16.

IMR operates one fixed station, Ingøy, related to the Barents Sea. The Ingøy station is situated in the coastal current along the Norwegian coast. Temperature and salinity is monitored 1-4 times a month. The observations were obtained in two periods, 19361944 and 1968-present.

### 1.6.2 Area coverage

Area surveys are conducted throughout the year. The number of vessels in each survey differs, not only between surveys but may also change from year to year for the same survey. However, most surveys are conducted with only one vessel. It is not possible to measure all ecosystem components during each survey. Effort is always put on measuring as many parameters as possible on each survey, but available time put restrictions on what is possible to accomplish. Also, an investigation should not take too long time in order to give a synoptic picture of the conditions. Therefore the surveys must focus on a specific set of parameters/species. Other measured parameters may therefore not have optimal coverage and thereby increased uncertainty, but will still give important information. An overview of the measured parameters/species on each main survey is given in Table 1.16. Specific considerations for the most important surveys are given in the following text.

## Norwegian/Russian winter survey

The survey is carried out during February-early March, and covers the main cod distribution area in the Barents Sea. The coverage is in some years limited by the ice distribution. Three vessels are normally applied, two Norwegian and one Russian. The main observations are made with bottom trawl, pelagic trawl, echo sounder and CTD. Plankton studies have been done in some years. Cod and haddock are the main targets for this survey. Swept area indices are calculated for cod, haddock, Greenland halibut, S. marinus and S. mentella. Acoustic observations are made for cod, haddock, capelin, redfish, polar cod and herring. The survey started in 1981.

## Lofoten survey

The main spawning grounds of North East Arctic cod are in the Lofoten area. Echosounder equipment was first used in 1935 to detect concentrations of spawning cod, and the first attempt to map such concentrations was made in 1938 (Sund, 1938). Later investigations have provided valuable information on the migratory patterns, the geographical distribution and the age composition and abundance of the stock.

The current time series of survey data starts in 1985. Due to the change in echo sounder equipment in 1990 results obtained earlier are not directly comparable with later results. The survey is designed as equidistant parallel acoustic transects covering 3 strata (North, South and Vestfjorden). In most surveys previous to 1990 the transects were not parallel, but more as parts of a zig-zag pattern across the spawning grounds aimed at mapping the distribution of cod. Trawl samples are not taken according to a proper trawl survey design. This is due to practical reasons. The spawning concentrations can be located with echosounder thus effectively reduce the number of trawl stations needed. The ability to properly sample the composition of the stock (age, sex, maturity stage etc.) is limited by the amount of fixed gear (gillnets and longlines) in the different areas.

## Norwegian coastal surveys

In 1985-2002 a Norwegian acoustic survey specially designed for saithe was conducted annually in October-November (Nedreaas 1998). The survey covered the near coastal banks from the Varangerfjord close to the Russian border and southwards to $62^{\circ} \mathrm{N}$. The whole area has been covered since 1992, and the major parts since 1988. The aim of conducting an acoustic survey targeting Northeast Arctic saithe was to support the stock assessment with fishery-independent data of the abundance of the youngest saithe. The survey mainly covered the grounds where the trawl fishery takes place, normally dominated by $3-5(6)$ year old fish. 2-year-old saithe, mainly inhabiting the fjords and more coastal areas, were also represented in the survey, although highly variable from year to year. In 1995-2002 a Norwegian acoustic survey for coastal cod was conducted along the coast and in the fjords from Varanger to Stad in September, just prior to the saithe survey described above. This survey covered coastal areas not included in the regular saithe survey. Autumn 2003 the saithe- and coastal cod surveys were combined. The survey now also covers 0 -group herring in fjords north of Lofoten.

## Joint ecosystem autumn survey

The survey is carried out from early August to early October, and covers the whole Barents Sea. Four or five vessels are normally applied, three Norwegian and one or two Russian. Most aspects of the ecosystem are covered, from physical and chemical oceanography, primary and secondary production, fish (both young and adult stages), sea mammals, benthos and birds. Many kinds of methods and gears are used, water sampling, plankton nets, pelagic and demersal trawls, grabs and sledges, acoustics, directs observations (birds and sea mammals). The survey has developed from joint surveys on 0-group, capelin and juvenile Greenland halibut, through general acoustic surveys including observations of physical oceanography and plankton, gradually developing into the ecosystem survey carried out in recent years. The predecessor of the survey dates back to 1972 and has been carried out every fall since. From 2003 these surveys were called "ecosystem surveys".

In 2009 not all components of the ecosystem were covered during the survey, and a further reduction will probably take place in 2010; the coverage of e.g. Greenland halibut will be less complete than in previous years. Also, the future of this ecosystem survey is still undetermined.

Associated with this survey Russia also covers parts of the Northern Kara Sea during autumn.

## Russian Autumn-winter trawl-acoustic survey

The survey is carried out in October-December, and cover the whole Barents Sea up to the continental slope. Two Russian vessels are usually used. The survey has developed from a young cod and haddock trawl survey, started in 1946. The current trawlacoustic time series of survey data starts in 1984, targeting both young and adult stages of bottom fish. The surveys include observations of physical oceanography and meso- and macro-zooplankton.

## Norwegian Greenland halibut survey

The survey is carried out in August, and cover the continental slope from 68 to $80^{\circ} \mathrm{N}$, in depths of $400-1500 \mathrm{~m}$ north of $70^{\circ} 30^{\prime} \mathrm{N}$, and $400-1000 \mathrm{~m}$ south of this latitude. This survey was run the first time in 1994, and is now part of the Norwegian Combined survey index for Greenland halibut. This survey will not be conducted in 2010, and its future design is being revised.

## Russian young herring survey

This survey is conducted in May and takes 2-3 weeks. It is including also observations of physical oceanography and plankton. In 1991-1995 it was joint survey, since 1996 the survey is carried out only by PINRO.

### 1.6.3 Other information sources

Large 3D hydrodynamic numeric models for the Barents Sea are run at both IMR and PINRO. These models have, through validation with observations, proved to be a useful tool for filling observation gaps in time and space. The hydrodynamic models have also proved useful for scenario testing, and for study of drift patterns of various planktonic organisms.

Sub-models for phytoplankton and zooplankton are now implemented in some of the hydrodynamic models. However, due to the present assumptions in these submodels care must be taken in the interpretation of the model results.

Satellites can be for several monitoring tasks. Ocean color specter can be used to identify and estimate the amount of phytoplankton in the skin ( $\sim 1 \mathrm{~m}$ ) layer. Several climate variables can be monitored (e.g. ice cover, cloud cover, heat radiation, sea surface temperature). Marine mammals, polar bears and seabirds can be traced with attached transmitters.

Aircraft surveys also are used for monitoring several physical parameters associated with the sea surface as well as observations of mammals at the surface and estimations of harp seal pup production in the White Sea.

Several international hindcast databases (e.g.. NCEP, ERA40) are available. They use a combination of numerical models and available observations to estimate several climate variables, covering the whole world.

Along the Norwegian coast ship-of-opportunity supply weekly the surface temperature along their path.

### 1.7 Main conclusions

State and expected situation in the ecosystem (section 1.2)

## Climate

- The air temperature was above the long-term mean during 2009.
- The sea temperature in the Barents Sea is still high, and about the same level as in 2008. There was an increase in the end of the year, with the highest December temperature in the Kola section. In 2010 the temperature is expected to further decrease, but still be higher than the long-term mean.
- Salinity in 2009 is still high , and at about the same levels as in 2008
- Inflow of Atlantic waters at the western entrance in 2009 was quite similar to 2008; moderate during winter followed by a strong decrease in spring. Data for second half of 2009 is not available.
- Oxygen levels were about normal in 2009
- Ice extent in 2008 was less than normal, but more than in 2008. In 2010 ice conditions is expected to be slightly less or around the long term mean.


## Plankton and northern shrimp

- The mesozooplankton biomass measured in August-September 2009 was less compared to 2008, and below the long-term mean.
- Abundance euphausiids (krill) in autumn and winter 2009 were close to the level in 2008 in the western and northwestern areas and increased in the centre, eastern and coastal areas. In total the abundance in 2010 is slightly above the long-term mean.
- The abundance of gelatinous zooplankton, caught by pelagic trawling, show a lower abundance in 2009 compared to 2008.
- The shrimp stock in the Barents Sea and Spitsbergen area in 2009 increased compared to both 2007 and 2008.


## Fish

- Capelin stock size is at around average level, with a slight decrease from last year. The survey estimate at age 1 of the 2008 year class is slightly below the long-term mean. 0-group estimates indicate that the 2009 year class is below average.
- For young herring there are indications that the year classes 2005-2009 are below average. Therefore the abundance of herring in the Barents Sea is believed to be at a relatively low level in 2010
- Blue whiting is still at a very low level, with a slight increase from 2008. The abundance is expected to remain low in 2009.
- The polar cod stock is presently at a high level, similar to 2008.


## Harp Seal

- The decrease in the harp seal pup production in the White Sea has become slower recently and even some slight increase has been observed, but it is still at a low level.


## Impact of fisheries on the ecosystem (section 1.3)

- The most widespread gear is trawl.
- The demersal fisheries are mixed, and currently have largest effect on coastal cod and redfish due to the poor condition of these stocks.
- The pelagic fisheries are less mixed, and are weakly linked to the demersal fisheries (however, by-catches of young pelagic stages of demersal species have been reported in some pelagic fisheries)
- Trawling has largest effect on hard bottom habitats; whereas the effects on other habitats are not clear and consistent.
- Work is currently going on exploring the possibility of using pelagic trawls when targeting demersal fish. The purpose is to avoid impact on bottom fauna and to reduce the mixture of other species. It will be mandatory to use sorting grids to avoid catches of undersized fish.
- Fishery induced mortality (lost gillnets, contact with active fishing gears, etc.) on fish is a potential problem but not quantified at present.


## Management improvement issues (section 1.4)

- Several methods, which take ecosystem information into account, are presently under development. These methods should in the future be valuable for the improvement of the stock assessment and advice.
- According to STOCOBAR simulations there is a low probability to expect any tendency of decline or increase in the fishable cod stock biomass in 2010 and 2011, based on predicted temperature and capelin stock size.
- The cod recruitment (age 3) in 2010 is expected to be low compared to the long-term mean. In 2011 and 2012 it is expected to increase slightly, but still be below the long-term mean.


### 1.8 Response to technical minutes

There were no specific comments from the review group to ecosystem consideration chapter (Chapter 1).

Table 1.1. 0 -group abundance indices (in millions) with $95 \%$ confidence limits, not corrected for catching efficiency

| Year | Capelin |  |  | Cod |  |  | Haddock |  |  | Herring |  |  | Redfish |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  |
| 1980 | 197278 | 131674 | 262883 | 72 | 38 | 105 | 59 | 38 | 81 | 4 | 1 | 8 | 277873 | 0 | 701273 |
| 1981 | 123870 | 71852 | 175888 | 48 | 33 | 64 | 15 | 7 | 22 | 3 | 0 | 8 | 153279 | 0 | 363283 |
| 1982 | 168128 | 35275 | 300982 | 651 | 466 | 835 | 649 | 486 | 812 | 202 | 0 | 506 | 106140 | 63753 | 148528 |
| 1983 | 100042 | 56325 | 143759 | 3924 | 1749 | 6099 | 1356 | 904 | 1809 | 40557 | 19526 | 61589 | 172392 | 33352 | 311432 |
| 1984 | 68051 | 43308 | 92794 | 5284 | 2889 | 7679 | 1295 | 937 | 1653 | 6313 | 1930 | 10697 | 83182 | 36137 | 130227 |
| 1985 | 21267 | 1638 | 40896 | 15484 | 7603 | 23365 | 695 | 397 | 992 | 7237 | 646 | 13827 | 412777 | 40510 | 785044 |
| 1986 | 11409 | 98 | 22721 | 2054 | 1509 | 2599 | 592 | 367 | 817 | 7 | 0 | 15 | 91621 | 0 | 184194 |
| 1987 | 1209 | 435 | 1983 | 167 | 86 | 249 | 126 | 76 | 176 | 2 | 0 | 5 | 23747 | 12740 | 34755 |
| 1988 | 19624 | 3821 | 35427 | 507 | 296 | 718 | 387 | 157 | 618 | 8686 | 3325 | 14048 | 107027 | 23378 | 190675 |
| 1989 | 251485 | 201110 | 301861 | 717 | 404 | 1030 | 173 | 117 | 228 | 4196 | 1396 | 6996 | 16092 | 7589 | 24595 |
| 1990 | 36475 | 24372 | 48578 | 6612 | 3573 | 9651 | 1148 | 847 | 1450 | 9508 | 0 | 23943 | 94790 | 52658 | 136922 |
| 1991 | 57390 | 24772 | 90007 | 10874 | 7860 | 13888 | 3857 | 2907 | 4807 | 81175 | 43230 | 119121 | 41499 | 0 | 83751 |
| 1992 | 970 | 105 | 1835 | 44583 | 24730 | 64437 | 1617 | 1150 | 2083 | 37183 | 21675 | 52690 | 13782 | 0 | 36494 |
| 1993 | 330 | 125 | 534 | 38015 | 15944 | 60086 | 1502 | 911 | 2092 | 61508 | 2885 | 120131 | 5458 | 0 | 13543 |
| 1994 | 5386 | 0 | 10915 | 21677 | 11980 | 31375 | 1695 | 825 | 2566 | 14884 | 0 | 31270 | 52258 | 0 | 121547 |
| 1995 | 862 | 0 | 1812 | 74930 | 38459 | 111401 | 472 | 269 | 675 | 1308 | 434 | 2182 | 11816 | 3386 | 20246 |
| 1996 | 44268 | 22447 | 66089 | 66047 | 42607 | 89488 | 1049 | 782 | 1316 | 57169 | 28040 | 86299 | 28 | 8 | 47 |
| 1997 | 54802 | 22682 | 86922 | 67061 | 49487 | 84634 | 600 | 420 | 780 | 45808 | 21160 | 70455 | 132 | 0 | 272 |
| 1998 | 33841 | 21406 | 46277 | 7050 | 4209 | 9890 | 5964 | 3800 | 8128 | 79492 | 44207 | 114778 | 755 | 23 | 1487 |
| 1999 | 85306 | 45266 | 125346 | 1289 | 135 | 2442 | 1137 | 368 | 1906 | 15931 | 1632 | 30229 | 46 | 14 | 79 |
| 2000 | 39813 | 1069 | 78556 | 26177 | 14287 | 38068 | 2907 | 1851 | 3962 | 49614 | 3246 | 95982 | 7530 | 0 | 16826 |
| 2001 | 33646 | 0 | 85901 | 908 | 152 | 1663 | 1706 | 1113 | 2299 | 844 | 177 | 1511 | 6 | 1 | 10 |
| 2002 | 19426 | 10648 | 28205 | 19157 | 11015 | 27300 | 1843 | 1276 | 2410 | 23354 | 12144 | 34564 | 130 | 20 | 241 |
| 2003 | 94902 | 41128 | 148676 | 17304 | 10225 | 24383 | 7910 | 3757 | 12063 | 28579 | 15504 | 41653 | 216 | 0 | 495 |
| 2004 | 16701 | 2541 | 30862 | 19157 | 13987 | 24328 | 19144 | 12649 | 25638 | 133350 | 94873 | 171826 | 849 | 0 | 1766 |
| 2005 | 41808 | 12316 | 71300 | 21532 | 14732 | 28331 | 33283 | 24377 | 42190 | 26332 | 1132 | 51532 | 12332 | 631 | 24034 |
| 2006 | 166400 | 102749 | 230050 | 7860 | 3658 | 12061 | 11421 | 7553 | 15289 | 66819 | 22759 | 110880 | 20864 | 10057 | 31671 |
| 2007 | 157913 | 87370 | 228456 | 9707 | 5887 | 13527 | 2826 | 1787 | 3866 | 22481 | 4556 | 40405 | 159159 | 44882 | 273436 |
| 2008 | 288799 | 178860 | 398738 | 52975 | 31839 | 74111 | 2742 | 830 | 4655 | 15915 | 4477 | 27353 | 9962 | 0 | 20828 |
| 2009 | 189767 | 113154 | 266379 | 54579 | 37311 | 71846 | 13040 | 7988 | 18093 | 18916 | 8249 | 29582 | 66671 | 29636 | 103706 |
| Mean | 77706 |  |  | 19880 |  |  | 4040 |  |  | 28579 |  |  | 64744 |  |  |

Table 1.1. (cont.). 0-group abundance indices (in millions) with $95 \%$ confidence limits, not corrected for catching efficiency.

| Year | Saithe |  | Gr halibut |  | Long rough dab |  | Polar cod (east) |  | Polar cod (west) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Abundance index | Confidence limit | Abundance index | Confidence limit | Abundance index | Confidence limit | Abundance index | Confidence limit | Abundance index | Confidence limit |
| 1980 | 3 | $0 \quad 6$ | 111 | $35 \quad 187$ | 1273 | 8831664 | 28958 | 978448132 | 9650 | 020622 |
| 1981 | 0 | 00 | 74 | $46 \quad 101$ | 556 | 300813 | 595 | 226963 | 5150 | 19568345 |
| 1982 | 143 | 0371 | 39 | $11 \quad 68$ | 1013 | 6981328 | 1435 | 1442725 | 1187 | 03298 |
| 1983 | 239 | $83 \quad 394$ | 41 | $22 \quad 59$ | 420 | 264577 | 1246 | 02501 | 9693 | 020851 |
| 1984 | 1339 | 4072271 | 31 | $18 \quad 45$ | 60 | $43 \quad 77$ | 127 | 0303 | 3182 | 7375628 |
| 1985 | 12 | 123 | 48 | $29 \quad 67$ | 265 | 110420 | 19220 | 498933451 | 809 | $0 \quad 1628$ |
| 1986 | 1 | $0 \quad 2$ | 112 | $60 \quad 164$ | 6846 | 49418752 | 12938 | $2355 \quad 23521$ | 2130 | 1804081 |
| 1987 | 1 | $0 \quad 1$ | 35 | $23 \quad 47$ | 804 | 4111197 | 7694 | 017552 | 74 | $31 \quad 117$ |
| 1988 | 17 | 430 | 8 | 313 | 205 | 113297 | 383 | $9 \quad 757$ | 4634 | 09889 |
| 1989 | 1 | 03 | 1 | 03 | 180 | 100260 | 199 | $0 \quad 423$ | 18056 | 218233931 |
| 1990 | 11 | 20 | 1 | $0 \quad 2$ | 55 | $26 \quad 84$ | 399 | $129 \quad 669$ | 31939 | 070847 |
| 1991 | 4 | 26 | 1 | $0 \quad 2$ | 90 | $49 \quad 131$ | 88292 | 39856136727 | 38709 | 0110568 |
| 1992 | 159 | $86 \quad 233$ | 9 | $0 \quad 17$ | 121 | $25 \quad 218$ | 7539 | $0 \quad 15873$ | 9978 | 159118365 |
| 1993 | 366 | 0913 | 4 | 27 | 56 | $25 \quad 87$ | 41207 | 096068 | 8254 | 135915148 |
| 1994 | 2 | $0 \quad 5$ | 39 | $0 \quad 93$ | 1696 | 10832309 | 267997 | 151917384078 | 5455 | $0 \quad 12032$ |
| 1995 | 148 | $68 \quad 229$ | 15 | $5 \quad 24$ | 229 | $39 \quad 419$ | 1 | $0 \quad 2$ | 25 | 49 |
| 1996 | 131 | $57 \quad 204$ | 6 | $3 \quad 9$ | 41 | 279 | 70134 | 4319697072 | 4902 | 012235 |
| 1997 | 78 | $37 \quad 120$ | 5 | $3 \quad 7$ | 97 | $44 \quad 150$ | 33580 | 1878848371 | 7593 | 62314563 |
| 1998 | 86 | $39 \quad 133$ | 8 | $3 \quad 12$ | 27 | $13 \quad 42$ | 11223 | 684915597 | 10311 | $0 \quad 23358$ |
| 1999 | 136 | $68 \quad 204$ | 14 | $8 \quad 21$ | 105 | 1210 | 129980 | 82936177023 | 2848 | 4075288 |
| 2000 | 206 | 111301 | 43 | $17 \quad 69$ | 233 | 120346 | 116121 | 67589164652 | 22740 | 1492430556 |
| 2001 | 20 | $0 \quad 46$ | 51 | $20 \quad 83$ | 162 | $78 \quad 246$ | 3697 | 6586736 | 13490 | 028796 |
| 2002 | 553 | 108998 | 51 | $0 \quad 112$ | 731 | 3421121 | 96954 | 57530136378 | 27753 | 418451322 |
| 2003 | 65 | $0 \quad 146$ | 13 | 034 | 78 | $45 \quad 110$ | 11211 | 610016323 | 1627 | 03643 |
| 2004 | 1395 | 8601930 | 70 | $28 \quad 113$ | 36 | $20 \quad 52$ | 37156 | 1904055271 | 367 | $125 \quad 610$ |
| 2005 | 55 | $36 \quad 73$ | 9 | 414 | 200 | 109292 | 6540 | 31969884 | 3216 | 12695162 |
| 2006 | 142 | $60 \quad 224$ | 11 | 120 | 710 | 437983 | 26016 | 999642036 | 2078 | 4643693 |
| 2007 | 51 | $6 \quad 96$ | 1 | 0 | 262 | $45 \quad 478$ | 25883 | 849443273 | 2532 | 05134 |
| 2008 | 45 | $22 \quad 69$ | 6 | $0 \quad 13$ | 956 | 4101502 | 6649 | 84512453 | 91 | $0 \quad 183$ |
| 2009 | 22 | $0 \quad 46$ | 7 | 410 | 115 | $51 \quad 179$ | 23570 | 966137479 | 21433 | 564237223 |
| Mean | 181 |  | 29 |  | 587 |  | 35898 |  | 8997 |  |

Table 1.2. 0-group abundance indices (in millions) with $95 \%$ confidence limits, corrected for catching efficiency.

| Year | Capelin |  |  | Cod |  |  | Haddock |  |  | Herring |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  | Abundance index | Confidence limit |  |
| 1980 | 740289 | 495187 | 985391 | 276 | 131 | 421 | 265 | 169 | 361 | 77 | 12 | 142 |
| 1981 | 477260 | 273493 | 681026 | 289 | 201 | 377 | 75 | 34 | 117 | 37 | 0 | 86 |
| 1982 | 599596 | 145299 | 1053893 | 3480 | 2540 | 4421 | 2927 | 2200 | 3655 | 2519 | 0 | 5992 |
| 1983 | 340200 | 191122 | 489278 | 19299 | 9538 | 29061 | 6217 | 3978 | 8456 | 195446 | 69415 | 321477 |
| 1984 | 275233 | 161408 | 389057 | 24326 | 14489 | 34164 | 5512 | 3981 | 7043 | 27354 | 3425 | 51284 |
| 1985 | 63771 | 5893 | 121648 | 66630 | 32914 | 100346 | 2457 | 1520 | 3393 | 20081 | 3933 | 36228 |
| 1986 | 41814 | 642 | 82986 | 10509 | 7719 | 13299 | 2579 | 1621 | 3537 | 93 | 27 | 160 |
| 1987 | 4032 | 1458 | 6607 | 1035 | 504 | 1565 | 708 | 432 | 984 | 49 | 0 | 111 |
| 1988 | 65127 | 12101 | 118153 | 2570 | 1519 | 3622 | 1661 | 630 | 2693 | 60782 | 20877 | 100687 |
| 1989 | 862394 | 690983 | 1033806 | 2775 | 1624 | 3925 | 650 | 448 | 852 | 17956 | 8252 | 27661 |
| 1990 | 115636 | 77306 | 153966 | 23593 | 13426 | 33759 | 3122 | 2318 | 3926 | 15172 | 0 | 36389 |
| 1991 | 169455 | 74078 | 264832 | 40631 | 29843 | 51419 | 13713 | 10530 | 16897 | 267644 | 107990 | 427299 |
| 1992 | 2337 | 250 | 4423 | 166276 | 92113 | 240438 | 4739 | 3217 | 6262 | 83909 | 48399 | 119419 |
| 1993 | 952 | 289 | 1616 | 133046 | 58312 | 207779 | 3785 | 2335 | 5236 | 291468 | 1429 | 581506 |
| 1994 | 13898 | 70 | 27725 | 70761 | 39933 | 101589 | 4470 | 2354 | 6586 | 103891 | 0 | 212765 |
| 1995 | 2869 | 0 | 6032 | 233885 | 114258 | 353512 | 1203 | 686 | 1720 | 11018 | 4409 | 17627 |
| 1996 | 136674 | 69801 | 203546 | 280916 | 188630 | 373203 | 2632 | 1999 | 3265 | 549608 | 256160 | 843055 |
| 1997 | 189372 | 80734 | 298011 | 294607 | 218967 | 370247 | 1983 | 1391 | 2575 | 463243 | 176669 | 749817 |
| 1998 | 113390 | 70516 | 156263 | 24951 | 15827 | 34076 | 14116 | 9524 | 18707 | 476065 | 277542 | 674589 |
| 1999 | 287760 | 143243 | 432278 | 4150 | 944 | 7355 | 2740 | 1018 | 4463 | 35932 | 13017 | 58848 |
| 2000 | 140837 | 6551 | 275123 | 108093 | 58416 | 157770 | 10906 | 6837 | 14975 | 469626 | 22507 | 916746 |
| 2001 | 90181 | 0 | 217345 | 4150 | 798 | 7502 | 4649 | 3189 | 6109 | 10008 | 2021 | 17996 |
| 2002 | 67130 | 36971 | 97288 | 76146 | 42253 | 110040 | 4381 | 2998 | 5764 | 151514 | 58954 | 244073 |
| 2003 | 340877 | 146178 | 535575 | 81977 | 47715 | 116240 | 30792 | 15352 | 46232 | 177676 | 52699 | 302653 |
| 2004 | 53950 | 11999 | 95900 | 65969 | 47743 | 84195 | 39303 | 26359 | 52246 | 773891 | 544964 | 1002819 |
| 2005 | 148466 | 51669 | 245263 | 72137 | 50662 | 93611 | 91606 | 67869 | 115343 | 125927 | 20407 | 231447 |
| 2006 | 515770 | 325776 | 705764 | 25061 | 11469 | 38653 | 28505 | 18754 | 38256 | 294649 | 102788 | 486511 |
| 2007 | 480069 | 272313 | 687825 | 42628 | 26652 | 58605 | 8401 | 5587 | 11214 | 144002 | 25099 | 262905 |
| 2008 | 995101 | 627202 | 1362999 | 234144 | 131081 | 337208 | 9864 | 1144 | 18585 | 201046 | 68778 | 333313 |
| 2009 | 673027 | 423386 | 922668 | 185457 | 123375 | 247540 | 33339 | 19707 | 46970 | 104233 | 31009 | 177458 |
| Mean | 266916 |  |  | 76659 |  |  | 11243 |  |  | 169164 |  |  |

Table 1.2 (cont.). 0-group abundance indices (in millions) with $95 \%$ confidence limits, corrected for catching efficiency.

| Year | Saithe |  | Polar cod (east) |  |  | Polar cod (west) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Abundance index | Confidence limit | Abundance index | Confidence limit |  | Abundance index | Confide | ce limit |
| 1980 | 21 | $0 \quad 47$ | 203226 | 69898 | 336554 | 82871 | 0 | 176632 |
| 1981 | 0 | $0 \quad 0$ | 4882 | 1842 | 7922 | 46155 | 17810 | 74500 |
| 1982 | 296 | 0699 | 1443 | 154 | 2731 | 10565 | 0 | 29314 |
| 1983 | 562 | 211912 | 1246 | 0 | 2501 | 87272 | 0 | 190005 |
| 1984 | 2577 | 7254430 | 871 | 0 | 2118 | 26316 | 6097 | 46534 |
| 1985 | 30 | 753 | 143257 | 39633 | 246881 | 6670 | 0 | 13613 |
| 1986 | 4 | 0 | 102869 | 16336 | 189403 | 18644 | 125 | 37164 |
| 1987 | 4 | $0 \quad 10$ | 64171 | 0 | 144389 | 631 | 265 | 996 |
| 1988 | 32 | $11 \quad 52$ | 2588 | 59 | 5117 | 41133 | 0 | 89068 |
| 1989 | 10 | 023 | 1391 | 0 | 2934 | 164058 | 15439 | 312678 |
| 1990 | 29 | 455 | 2862 | 879 | 4846 | 246819 | 0 | 545410 |
| 1991 | 9 | 414 | 823828 | 366924 | 1280732 | 281434 | 0 | 799822 |
| 1992 | 326 | 156495 | 49757 | 0 | 104634 | 80747 | 12984 | 148509 |
| 1993 | 1033 | 02512 | 297397 | 0 | 690030 | 70019 | 12321 | 127716 |
| 1994 | 7 | $1 \quad 12$ | 2139223 | 1230225 | 3048220 | 49237 | 0 | 109432 |
| 1995 | 415 | $196 \quad 634$ | 6 | 0 | 14 | 195 | 0 | 390 |
| 1996 | 430 | $180 \quad 679$ | 588020 | 368361 | 807678 | 46671 | 0 | 116324 |
| 1997 | 341 | 162521 | 297828 | 164107 | 431550 | 62084 | 6037 | 118131 |
| 1998 | 182 | $91 \quad 272$ | 96874 | 59118 | 134630 | 95609 | 0 | 220926 |
| 1999 | 275 | 139411 | 1154149 | 728616 | 1579682 | 24015 | 3768 | 44262 |
| 2000 | 851 | 4461256 | 916625 | 530966 | 1302284 | 190661 | 133249 | 248072 |
| 2001 | 47 | $0 \quad 106$ | 29087 | 5648 | 52526 | 119023 | 0 | 252146 |
| 2002 | 2112 | 1344090 | 829216 | 496352 | 1162079 | 215572 | 36403 | 394741 |
| 2003 | 286 | 0631 | 82315 | 42707 | 121923 | 12998 | 0 | 30565 |
| 2004 | 4779 | $2810 \quad 6749$ | 290686 | 147492 | 433879 | 2892 | 989 | 4796 |
| 2005 | 176 | $115 \quad 237$ | 44663 | 22890 | 66436 | 25970 | 9987 | 41953 |
| 2006 | 280 | 116443 | 182713 | 73645 | 291781 | 15965 | 3414 | 28517 |
| 2007 | 286 | 568 | 191111 | 57403 | 324819 | 22803 | 0 | 46521 |
| 2008 | 142 | $68 \quad 216$ | 42657 | 5936 | 79378 | 619 | 25 | 1212 |
| 2009 | 62 | $0 \quad 132$ | 168990 | 70509 | 267471 | 154687 | 37022 | 272351 |
| Mean | 520 |  | 291798 |  |  | 73411 |  |  |

Table 1.3. The North-east arctic cod stock's consumption of various prey species in 1984-2009 (1000 tonnes), based on Norwegian consumption calculations.

| Year | Other | Amphipods | Krill | Shrimp | Capelin | Herring | Polar cod | Cod | Haddock | Redfish | G. halibut | $\begin{array}{r} \text { Blue } \\ \text { whiting } \end{array}$ | Long rough <br> dab | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 479 | 27 | 113 | 436 | 722 | 78 | 15 | 22 | 50 | 364 | 0 | 0 | 24 | 2330 |
| 1985 | 1112 | 170 | 58 | 156 | 1621 | 183 | 3 | 32 | 47 | 225 | 0 | 1 | 41 | 3649 |
| 1986 | 606 | 1236 | 111 | 142 | 837 | 133 | 141 | 83 | 111 | 315 | 0 | 0 | 55 | 3769 |
| 1987 | 671 | 1085 | 67 | 191 | 229 | 32 | 206 | 25 | 4 | 324 | 1 | 0 | 9 | 2844 |
| 1988 | 401 | 1237 | 318 | 129 | 339 | 8 | 92 | 9 | 3 | 223 | 0 | 4 | 5 | 2769 |
| 1989 | 656 | 800 | 241 | 131 | 572 | 3 | 32 | 8 | 10 | 228 | 0 | 0 | 57 | 2739 |
| 1990 | 1343 | 137 | 85 | 195 | 1609 | 7 | 6 | 19 | 15 | 243 | 0 | 87 | 95 | 3842 |
| 1991 | 760 | 65 | 76 | 188 | 2891 | 8 | 12 | 26 | 20 | 312 | 7 | 10 | 270 | 4646 |
| 1992 | 907 | 102 | 158 | 373 | 2457 | 331 | 97 | 55 | 106 | 188 | 20 | 2 | 93 | 4889 |
| 1993 | 751 | 253 | 714 | 315 | 3033 | 163 | 278 | 285 | 71 | 100 | 2 | 2 | 26 | 5994 |
| 1994 | 625 | 563 | 704 | 518 | 1085 | 147 | 582 | 224 | 49 | 79 | 0 | 1 | 39 | 4614 |
| 1995 | 845 | 982 | 516 | 362 | 628 | 115 | 254 | 371 | 116 | 193 | 1 | 0 | 34 | 4417 |
| 1996 | 599 | 631 | 1158 | 341 | 538 | 47 | 104 | 536 | 69 | 97 | 0 | 10 | 34 | 4164 |
| 1997 | 443 | 382 | 519 | 316 | 907 | 5 | 113 | 338 | 41 | 36 | 0 | 33 | 14 | 3146 |
| 1998 | 411 | 363 | 455 | 325 | 714 | 86 | 151 | 155 | 33 | 9 | 0 | 13 | 15 | 2730 |
| 1999 | 378 | 145 | 271 | 250 | 1720 | 128 | 220 | 62 | 26 | 16 | 1 | 31 | 7 | 3255 |
| 2000 | 385 | 167 | 464 | 450 | 1727 | 53 | 194 | 76 | 51 | 8 | 0 | 38 | 18 | 3633 |
| 2001 | 685 | 172 | 376 | 277 | 1722 | 71 | 250 | 66 | 49 | 6 | 1 | 151 | 29 | 3853 |
| 2002 | 362 | 96 | 261 | 232 | 1934 | 86 | 270 | 108 | 123 | 1 | 0 | 224 | 15 | 3713 |
| 2003 | 548 | 282 | 529 | 240 | 2157 | 214 | 272 | 114 | 168 | 3 | 0 | 74 | 48 | 4649 |
| 2004 | 671 | 679 | 318 | 247 | 1296 | 196 | 338 | 122 | 193 | 3 | 12 | 74 | 62 | 4212 |
| 2005 | 685 | 411 | 521 | 264 | 1238 | 187 | 354 | 116 | 342 | 2 | 3 | 111 | 46 | 4282 |
| 2006 | 780 | 169 | 957 | 313 | 1511 | 201 | 118 | 70 | 361 | 15 | 1 | 122 | 104 | 4721 |
| 2007 | 1141 | 293 | 935 | 373 | 1881 | 272 | 228 | 94 | 355 | 40 | 1 | 39 | 61 | 5712 |
| 2008 | 1384 | 146 | 787 | 316 | 2443 | 102 | 476 | 182 | 303 | 55 | 11 | 29 | 90 | 6325 |
| 2009 | 1250 | 159 | 427 | 211 | 2762 | 219 | 478 | 175 | 295 | 35 | 1 | 5 | 91 | 6109 |

Table 1.4. The North-east arctic COD stock's consumption of various prey species in 1984-2009 (1000 tonnes), based on Russian consumption calculations.

| Year |  |  |  |  |  |  | $\begin{aligned} & 0 \\ & 0 \\ & \hline \end{aligned}$ |  |  | $\begin{aligned} & \vec{Z} \\ & 0 \\ & 0 \\ & \vdots \\ & \vdots \\ & 3 \\ & 0 \\ & Z \\ & \hline \end{aligned}$ |  |  |  | $\begin{aligned} & \frac{\pi}{5} \\ & \ddot{む} \\ & 0 \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { ర్ } \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & \hline \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 93 | 31 | 351 | 33 | 592 | 17 | 13 | 50 | 5 | 1 | 195 | 51 | 0 | 269 | 286 | 1988 |
| 1985 | 30 | 432 | 202 | 24 | 989 | 0 | 98 | 34 | 18 | 15 | 97 | 23 | 0 | 519 | 198 | 2679 |
| 1986 | 57 | 860 | 148 | 47 | 807 | 159 | 28 | 103 | 3 | 27 | 158 | 24 | 1 | 372 | 170 | 2962 |
| 1987 | 69 | 508 | 201 | 8 | 162 | 105 | 27 | 2 | 10 | 15 | 118 | 6 | 0 | 268 | 188 | 1686 |
| 1988 | 209 | 169 | 118 | 19 | 292 | 0 | 20 | 93 | 0 | 0 | 127 | 20 | 0 | 239 | 242 | 1545 |
| 1989 | 167 | 290 | 104 | 4 | 680 | 34 | 34 | 2 | 0 | 0 | 158 | 56 | 0 | 201 | 248 | 1977 |
| 1990 | 101 | 30 | 270 | 64 | 1254 | 8 | 21 | 16 | 39 | 15 | 232 | 79 | 0 | 101 | 167 | 2396 |
| 1991 | 54 | 83 | 287 | 28 | 3286 | 44 | 52 | 22 | 7 | 6 | 144 | 46 | 6 | 132 | 158 | 4354 |
| 1992 | 213 | 38 | 263 | 374 | 2021 | 191 | 84 | 38 | 0 | 77 | 121 | 44 | 1 | 295 | 418 | 4175 |
| 1993 | 186 | 177 | 223 | 177 | 2791 | 171 | 147 | 153 | 4 | 25 | 41 | 48 | 5 | 160 | 384 | 4691 |
| 1994 | 362 | 298 | 472 | 105 | 1303 | 492 | 391 | 72 | 1 | 2 | 56 | 40 | 0 | 100 | 353 | 4046 |
| 1995 | 396 | 465 | 550 | 192 | 691 | 203 | 557 | 130 | 0 | 1 | 113 | 53 | 3 | 169 | 356 | 3878 |
| 1996 | 973 | 361 | 200 | 76 | 478 | 79 | 473 | 60 | 9 | 37 | 71 | 47 | 0 | 470 | 175 | 3509 |
| 1997 | 386 | 85 | 207 | 54 | 523 | 110 | 409 | 35 | 3 | 0 | 37 | 33 | 2 | 97 | 399 | 2380 |
| 1998 | 615 | 205 | 265 | 70 | 852 | 129 | 129 | 23 | 23 | 18 | 15 | 19 | 0 | 53 | 226 | 2641 |
| 1999 | 454 | 77 | 242 | 74 | 1402 | 165 | 48 | 14 | 25 | 1 | 13 | 8 | 0 | 58 | 107 | 2688 |
| 2000 | 413 | 111 | 367 | 48 | 1662 | 157 | 57 | 29 | 26 | 8 | 4 | 20 | 0 | 36 | 181 | 3119 |
| 2001 | 418 | 74 | 308 | 88 | 1433 | 140 | 59 | 49 | 137 | 29 | 4 | 31 | 2 | 145 | 190 | 3106 |
| 2002 | 309 | 45 | 198 | 55 | 2330 | 281 | 100 | 77 | 102 | 3 | 4 | 17 | 0 | 44 | 170 | 3734 |
| 2003 | 240 | 140 | 213 | 144 | 1155 | 204 | 127 | 323 | 26 | 5 | 1 | 38 | 0 | 87 | 270 | 2974 |
| 2004 | 350 | 378 | 243 | 122 | 1046 | 350 | 83 | 151 | 48 | 20 | 7 | 58 | 15 | 179 | 267 | 3317 |
| 2005 | 543 | 135 | 226 | 170 | 962 | 318 | 114 | 275 | 68 | 42 | 7 | 45 | 2 | 162 | 203 | 3272 |
| 2006 | 887 | 62 | 210 | 239 | 1186 | 108 | 95 | 268 | 104 | 86 | 17 | 95 | 1 | 92 | 333 | 3781 |
| 2007 | 860 | 153 | 280 | 259 | 1408 | 239 | 73 | 319 | 33 | 21 | 22 | 65 | 1 | 194 | 376 | 4304 |
| 2008 | 617 | 36 | 229 | 102 | 2324 | 498 | 138 | 333 | 16 | 16 | 42 | 109 | 13 | 301 | 416 | 5191 |
| 2009 | 511 | 105 | 199 | 158 | 2380 | 575 | 115 | 306 | 7 | 82 | 27 | 185 | 0 | 133 | 510 | 5293 |
| Mean | 366 | 206 | 253 | 105 | 1308 | 184 | 134 | 114 | 27 | 21 | 70 | 48 | 2 | 188 | 269 | 3296 |

Table 1.5. Consumption per cod by cod age group ( $\mathrm{kg} / \mathrm{year}$ ), based on Norwegian consumption calculations.

| Year/Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0.247 | 0.814 | 1.684 | 2.513 | 3.948 | 5.203 | 7.973 | 8.486 | 9.139 | 9.867 | 9.941 |
| 1985 | 0.304 | 0.761 | 1.829 | 3.101 | 4.671 | 7.357 | 11.172 | 11.892 | 12.416 | 13.660 | 13.773 |
| 1986 | 0.160 | 0.488 | 1.347 | 3.158 | 5.604 | 6.834 | 10.989 | 11.899 | 12.701 | 13.461 | 13.694 |
| 1987 | 0.219 | 0.601 | 1.275 | 2.055 | 3.537 | 5.457 | 7.044 | 8.111 | 8.922 | 9.343 | 9.295 |
| 1988 | 0.164 | 0.703 | 1.149 | 2.148 | 3.744 | 5.875 | 10.096 | 11.218 | 12.570 | 13.122 | 13.345 |
| 1989 | 0.223 | 0.716 | 1.606 | 2.705 | 3.973 | 5.601 | 7.648 | 8.464 | 9.559 | 10.156 | 10.599 |
| 1990 | 0.363 | 0.905 | 1.889 | 3.027 | 4.156 | 5.323 | 6.249 | 6.666 | 6.698 | 7.039 | 7.675 |
| 1991 | 0.293 | 0.969 | 2.168 | 3.500 | 5.281 | 7.026 | 9.392 | 10.154 | 11.200 | 12.239 | 11.886 |
| 1992 | 0.215 | 0.663 | 2.095 | 3.133 | 4.142 | 5.093 | 7.832 | 8.965 | 9.352 | 10.071 | 10.115 |
| 1993 | 0.112 | 0.528 | 1.546 | 3.044 | 4.809 | 6.285 | 9.421 | 11.239 | 11.763 | 12.253 | 12.876 |
| 1994 | 0.130 | 0.408 | 0.922 | 2.521 | 3.504 | 4.511 | 6.396 | 8.846 | 9.672 | 9.977 | 10.176 |
| 1995 | 0.103 | 0.296 | 0.921 | 1.840 | 3.361 | 5.252 | 7.697 | 10.405 | 12.333 | 12.734 | 13.180 |
| 1996 | 0.108 | 0.356 | 0.929 | 1.847 | 3.068 | 4.429 | 7.381 | 11.143 | 14.702 | 14.876 | 15.265 |
| 1997 | 0.140 | 0.319 | 0.940 | 1.768 | 2.710 | 3.536 | 5.253 | 8.149 | 12.582 | 13.484 | 13.091 |
| 1998 | 0.117 | 0.397 | 0.983 | 1.942 | 2.923 | 4.186 | 5.746 | 8.061 | 11.339 | 11.850 | 11.903 |
| 1999 | 0.163 | 0.505 | 1.093 | 2.717 | 3.717 | 5.442 | 6.965 | 9.179 | 11.004 | 12.007 | 12.109 |
| 2000 | 0.170 | 0.499 | 1.243 | 2.461 | 4.252 | 5.651 | 7.951 | 9.364 | 12.485 | 13.258 | 13.299 |
| 2001 | 0.171 | 0.456 | 1.309 | 2.439 | 3.682 | 5.294 | 7.523 | 11.085 | 13.422 | 14.117 | 14.434 |
| 2002 | 0.199 | 0.551 | 1.167 | 2.441 | 3.380 | 4.719 | 6.357 | 9.039 | 10.224 | 11.538 | 10.921 |
| 2003 | 0.207 | 0.653 | 1.312 | 2.390 | 3.995 | 5.946 | 8.411 | 10.405 | 12.786 | 13.397 | 14.343 |
| 2004 | 0.194 | 0.474 | 1.280 | 2.529 | 3.882 | 5.588 | 7.323 | 11.213 | 16.665 | 18.557 | 17.980 |
| 2005 | 0.194 | 0.653 | 1.376 | 2.592 | 3.918 | 5.588 | 7.182 | 9.771 | 13.090 | 14.012 | 14.784 |
| 2006 | 0.181 | 0.595 | 1.589 | 2.796 | 4.185 | 5.870 | 7.482 | 11.255 | 13.695 | 14.692 | 15.613 |
| 2007 | 0.213 | 0.621 | 1.742 | 3.178 | 4.704 | 6.231 | 7.802 | 9.621 | 12.636 | 13.223 | 13.808 |
| 2008 | 0.189 | 0.665 | 1.460 | 3.047 | 4.336 | 6.667 | 8.135 | 10.842 | 14.166 | 14.673 | 14.883 |
| 2009 | 0.182 | 0.586 | 1.414 | 2.888 | 4.568 | 5.789 | 8.074 | 10.195 | 12.252 | 13.203 | 13.286 |
| Average | 0.191 | 0.584 | 1.396 | 2.606 | 4.001 | 5.564 | 7.822 | 9.835 | 11.821 | 12.589 | 12.780 |

Table 1.6. Consumption per cod by cod age group (kg/year), based on Russian consumption calculations.

| Year/Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0.262 | 0.893 | 1.612 | 2.748 | 3.848 | 5.486 | 6.990 | 8.563 | 10.574 | 13.166 | 12.437 | 14.282 | 15.272 |
| 1985 | 0.295 | 0.752 | 1.656 | 2.683 | 4.264 | 6.601 | 8.242 | 9.743 | 10.975 | 14.447 | 16.499 | 16.061 | 17.343 |
| 1986 | 0.179 | 0.515 | 1.461 | 3.467 | 4.956 | 5.913 | 6.477 | 8.156 | 9.766 | 11.455 | 12.500 | 13.577 | 14.772 |
| 1987 | 0.145 | 0.431 | 0.844 | 1.561 | 3.078 | 4.346 | 7.279 | 9.683 | 12.703 | 14.482 | 15.014 | 15.115 | 16.377 |
| 1988 | 0.183 | 0.704 | 1.075 | 1.627 | 2.392 | 4.387 | 8.208 | 9.978 | 10.867 | 16.536 | 14.352 | 15.765 | 12.361 |
| 1989 | 0.282 | 0.910 | 1.468 | 2.207 | 3.244 | 4.799 | 6.581 | 8.725 | 11.134 | 15.799 | 15.950 | 17.909 | 14.023 |
| 1990 | 0.288 | 1.007 | 1.696 | 2.694 | 3.278 | 3.833 | 5.584 | 6.871 | 10.716 | 11.428 | 12.660 | 15.053 | 16.064 |
| 1991 | 0.241 | 0.936 | 2.670 | 4.473 | 6.038 | 7.846 | 9.590 | 11.542 | 14.970 | 19.294 | 17.509 | 20.109 | 22.109 |
| 1992 | 0.178 | 0.969 | 2.475 | 2.866 | 3.995 | 5.138 | 6.724 | 7.414 | 8.754 | 12.304 | 13.518 | 13.744 | 14.908 |
| 1993 | 0.133 | 0.476 | 1.512 | 2.865 | 3.944 | 5.108 | 7.372 | 8.945 | 10.343 | 11.600 | 14.067 | 14.893 | 15.922 |
| 1994 | 0.180 | 0.512 | 1.212 | 2.402 | 3.517 | 5.359 | 7.560 | 10.001 | 11.818 | 12.896 | 13.554 | 15.902 | 16.806 |
| 1995 | 0.194 | 0.497 | 0.962 | 1.819 | 3.204 | 4.847 | 7.332 | 9.688 | 13.835 | 15.247 | 16.960 | 18.230 | 19.202 |
| 1996 | 0.170 | 0.498 | 1.028 | 1.916 | 3.075 | 4.189 | 6.987 | 10.212 | 12.185 | 13.426 | 14.581 | 16.214 | 16.876 |
| 1997 | 0.119 | 0.341 | 0.992 | 1.908 | 2.668 | 3.503 | 4.954 | 7.980 | 12.174 | 21.523 | 20.666 | 21.822 | 24.237 |
| 1998 | 0.232 | 0.528 | 1.081 | 2.016 | 2.823 | 4.089 | 5.469 | 7.346 | 9.586 | 13.012 | 14.455 | 15.579 | 16.201 |
| 1999 | 0.261 | 0.431 | 1.128 | 2.490 | 3.676 | 5.222 | 6.398 | 8.220 | 9.194 | 13.364 | 15.325 | 16.918 | 17.567 |
| 2000 | 0.186 | 0.545 | 1.288 | 2.551 | 4.387 | 6.559 | 8.833 | 10.483 | 11.522 | 15.132 | 17.155 | 19.717 | 20.514 |
| 2001 | 0.150 | 0.413 | 1.163 | 2.110 | 3.430 | 5.571 | 6.835 | 10.233 | 12.457 | 15.130 | 17.374 | 19.322 | 20.559 |
| 2002 | 0.252 | 0.677 | 1.303 | 2.699 | 3.847 | 5.591 | 7.846 | 10.796 | 13.238 | 18.787 | 17.902 | 20.202 | 21.207 |
| 2003 | 0.228 | 0.618 | 1.296 | 2.028 | 3.547 | 4.716 | 6.684 | 8.905 | 13.418 | 14.492 | 19.540 | 19.239 | 20.036 |
| 2004 | 0.250 | 0.654 | 1.412 | 2.567 | 3.857 | 5.660 | 7.730 | 11.126 | 15.907 | 20.770 | 21.687 | 24.852 | 25.892 |
| 2005 | 0.255 | 0.687 | 1.514 | 2.504 | 3.896 | 5.264 | 7.192 | 9.395 | 13.163 | 15.981 | 22.656 | 23.387 | 24.181 |
| 2006 | 0.354 | 0.921 | 1.833 | 2.763 | 3.986 | 5.317 | 7.396 | 10.202 | 12.762 | 16.462 | 21.563 | 25.940 | 26.875 |
| 2007 | 0.234 | 0.666 | 1.803 | 3.018 | 4.295 | 5.810 | 7.444 | 9.017 | 11.754 | 15.961 | 20.903 | 25.154 | 26.064 |
| 2008 | 0.223 | 0.706 | 1.641 | 2.881 | 4.071 | 6.006 | 7.705 | 10.317 | 13.471 | 17.596 | 22.968 | 27.431 | 27.328 |
| 2009 | 0.217 | 0.627 | 1.503 | 2.542 | 4.266 | 5.530 | 7.617 | 10.986 | 13.258 | 15.637 | 21.532 | 25.632 | 25.586 |

Table 1.7. Capelin stock history from 1973-present. M output biomass is the estimated biomass of capelin removed from the stock by natural mortality.

| Year | Total stock number, billions (Oct. 1) | Total stock biomass in 1000 tonnes (Oct. 1) | Maturing biomass in 1000 tonnes (Oct. 1) | M output biomass (MOB) during year (1000 tonnes) |
| :---: | :---: | :---: | :---: | :---: |
| 1973 | 961 | 5144 | 1350 | 5504 |
| 1974 | 1029 | 5733 | 907 | 4542 |
| 1975 | 921 | 7806 | 2916 | 4669 |
| 1976 | 696 | 6417 | 3200 | 5633 |
| 1977 | 681 | 4796 | 2676 | 4174 |
| 1978 | 561 | 4247 | 1402 | 3782 |
| 1979 | 464 | 4162 | 1227 | 5723 |
| 1980 | 654 | 6715 | 3913 | 5708 |
| 1981 | 660 | 3895 | 1551 | 5658 |
| 1982 | 735 | 3779 | 1591 | 3729 |
| 1983 | 754 | 4230 | 1329 | 3884 |
| 1984 | 393 | 2964 | 1208 | 3051 |
| 1985 | 109 | 860 | 285 | 1975 |
| 1986 | 14 | 120 | 65 | 681 |
| 1987 | 39 | 101 | 17 | 200 |
| 1988 | 50 | 428 | 200 | 80 |
| 1989 | 209 | 864 | 175 | 537 |
| 1990 | 894 | 5831 | 2617 | 415 |
| 1991 | 1016 | 7287 | 2248 | 3307 |
| 1992 | 678 | 5150 | 2228 | 7745 |
| 1993 | 75 | 796 | 330 | 4631 |
| 1994 | 28 | 200 | 94 | 982 |
| 1995 | 17 | 193 | 118 | 163 |
| 1996 | 96 | 503 | 248 | 261 |
| 1997 | 140 | 911 | 312 | 828 |
| 1998 | 263 | 2056 | 931 | 915 |
| 1999 | 285 | 2776 | 1718 | 2070 |
| 2000 | 595 | 4273 | 2099 | 2464 |
| 2001 | 364 | 3630 | 2019 | 3906 |
| 2002 | 201 | 2210 | 1290 | 2939 |
| 2003 | 104 | 533 | 280 | 3195 |
| 2004 | 82 | 628 | 293 | 812 |
| 2005 | 42 | 324 | 174 | 817 |
| 2006 | 88 | 787 | 437 | 733 |
| 2007 | 280 | 1885 | 836 | 2033 |
| 2008 | 570 | 4426 | 2468 | 3285 |
| 2009 | 352 | 3756 | 2322 | * |

[^1]Table 1.8. Diet composition of main fish species in 2005, \% by weight (Data from Dolgov, WD 28 and WD 29, AFWG 2006)

| Prey species | Predators species |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Cod <br> (3+) | haddock | Greenland halibut | Thorny skate | Long rough dab | Saithe | Blue whiting |
| Euphausiidae | 5,2 | 21,7 | 0,4 | 0,8 | 0,1 | 24,4 | 44,4 |
| Hyperiidae | 4,1 | 0,2 | 3,8 | 0 | 0 | 0,3 | 18,2 |
| Cephalopoda | 0 | 0 | 2,1 | 0 | 0 | 0 | 0 |
| Pandalus borealis | 4,6 | 1,2 | 1,4 | 15,8 | 1,4 | 0,2 | 1,4 |
| Echinodermata | 0 | 24,1 | 0 | 0 | 4,7 | 0 | 0 |
| Mollusca | 0 | 7,9 | 0 | 0 | 3,6 | 0 | 0 |
| Polychaeta | 0 | 9,2 | 0 | 4,2 | 2,9 | 0 | 0 |
| Cod | 4,5 | 0,4 | 0,2 | 0 | 0,5 | 0,3 | 1,7 |
| Herring | 8,9 | 0,2 | 1,3 | 0,5 | 0,6 | 3,0 | 0 |
| Capelin | 11,6 | 2,1 | 8,7 | 30,8 | 17,5 | 54,9 | 0,9 |
| Haddock | 10,7 | 0,2 | 6,6 | 0,6 | 10,1 | 8,0 | 0 |
| Polar cod | 10,4 | 0 | 16,5 | 0 | 11,6 | 0,2 | 4,7 |
| Blue whiting | 4,8 | 0 | 2,6 | 0 | 0 | 0 | 0 |
| Greenland halibut | 0,2 | 0 | 1,4 | 0 | 0 | 0 | 0 |
| Redfish | 0,4 | 0 | 0,1 | 0 | 0 | 0 | 0 |
| Long rough dab | 1,8 | 0,1 | 4,8 | 2,9 | 0 | 0 | 0 |
| Other fish | 23,6 | 3,7 | 31,9 | 31,6 | 7,8 | 7,0 | 25,5 |
| Other food | 8,9 | 22,4 | 0,3 | 7,9 | 7,2 | 0 | 2,6 |
| Fishery waste | 0 | 4,1 | 17,7 | 4,9 | 31,4 | 0,9 | 0 |
| Undetermined | 0 | 2,4 | 0,2 | 1,4 | 0,7 | 0,5 | 0,3 |
| Total number of stomachs | 12209 | 7078 | 5223 | 432 | 2221 | 776 | 575 |
| Percentage of empty stomachs | 28,9 | 21,1 | 71,5 | 23,8 | 54,4 | 34,1 | 33,4 |
| Average filling degree | 1,7 | 1,6 | 0,7 | 1,9 | 1,1 | 1,6 | 1,7 |
| Mean index of stomach fullness | 213,8 | 110,5 | 84,4 | 182,7 | 139,0 | 116,3 | 111,2 |

Table 1.9. Annual consumption by minke whale and harp seal (thousand tonnes). The figures for minke whales are based on data from 1992-1995, while the figures for harp seals are based on data for 1990-1996.

| Prey | Minke <br> CONSUMPTION | Whale | HARP SEAL CONSUMPTION <br> (LOW CAPELIN STOCK) | HARP SEAL CONSUMPTION <br> (HiGH CAPELIN STOCK) |
| :--- | :--- | :--- | :--- | :--- |
| Capelin | 142 | 23 | 812 |  |
| Herring | 633 | 394 | 213 |  |
| Cod | 256 | 298 | 101 |  |
| Haddock | 128 | 47 | 1 |  |
| Krill | 602 | 550 | 605 |  |
| Amphipods | 0 | 304 | $313^{2}$ |  |
| Shrimp | 0 | 1 | 1 |  |
| Polar cod | 1 | 880 | 608 |  |
| Other fish | 55 | 622 | 406 |  |
| Other crustaceans | 0 | 356 | 312 |  |
| Total | 1817 | 3491 | 3371 |  |

${ }^{1}$ the prey species is included in the relevant 'other' group for this predator.
${ }^{2}$ only Parathemisto

Table 1.10. Description of the fisheries by gears. The gears are abbreviated as: trawl roundfish (TR), trawl shrimp (TS), longline (LL), gillnet (GN), handline (HL), purse seine (PS), Danish seine (DS) and trawl pelagic (TP). The regulations are abbreviated as: Quota (Q), mesh size (MS), sorting grid (SG), minimum catching size (MCS), maximum by-catch of undersized fish (MBU), maximum by-catch of non-target species (MBN), maximum as by-catch (MB), closure of areas (C), restrictions in season (RS), restrictions in area (RA), restriction in gear (RG), maximum by-catch per haul (MBH), as by-catch by maximum per boat at landing (MBL), number of effective fishing days (ED), number of vessels (EF).

| Species | Directed <br> FISHERY BY GEAR | TYPE OF FISHERY | LANDINGS IN 2008A (TONNES) | As BY-CATCH in fleet(s) | Location | AGreements and REGULATIONS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Capelin | PS, TP | seasonal | $10000{ }^{\text {B }}$ | TR, TS | Northern coastal areas to south of $74{ }^{\circ} \mathrm{N}$ | Bilateral agreement, Norway and Russia |
| Coastal cod | $\begin{aligned} & \text { GN, LL, HL, } \\ & \text { DS } \end{aligned}$ | all year | $25777 C$ | $\begin{aligned} & \text { TS, PS, DS, } \\ & \text { TP } \end{aligned}$ | Norwegian coast (inside 12 naut.miles) north of $62^{\circ} \mathrm{N}$ | $\begin{aligned} & \text { Q, MS, MCS, MBU, } \\ & \text { MBN, C, RS, RA } \end{aligned}$ |
| NEA Cod | $\begin{aligned} & \text { TR, GN, LL, } \\ & \text { HL } \end{aligned}$ | all year | 464 171 ${ }^{\text {C }}$ | $\begin{aligned} & \text { TS, PS, TP, } \\ & \text { DS } \end{aligned}$ | North of $62^{\circ} \mathrm{N}$, Barents Sea, Svalbard | $\begin{aligned} & \text { Q, MS, SG, MCS, MBU, } \\ & \text { MBN, C, RS, RA } \end{aligned}$ |
| Wolffish | LL | all year | 11355 | TR, (GN), <br> (HL) | North of $62^{\circ} \mathrm{N}$, Barents Sea, Svalbard | Q, MB |
| Haddock | $\begin{aligned} & \text { TR, GN, LL, } \\ & \text { HL } \end{aligned}$ | all year | 155604 | $\begin{aligned} & \text { TS, PS, TP, } \\ & \text { DS } \end{aligned}$ | North of $62^{\circ} \mathrm{N}$, Barents Sea, Svalbard | Q, MS, SG, MCS, MBU, MBN, C, RS, RA |
| Saithe | PS, TR, GN | seasonal | 183443 | $\begin{aligned} & \text { TS, LL, HL, } \\ & \text { DS, TP } \end{aligned}$ | Coastal areas north of $62^{\circ} \mathrm{N}$, southern Barents Sea | $\begin{aligned} & \text { Q, MS, SG, MCS, MBU, } \\ & \text { MBN, C, RS, RA } \end{aligned}$ |
| Greenland halibut | LL, GN | seasonal | 13144 | TR | Deep shelf and at the continental slope | Q, MS, RS, RG, MBH, MBL |
| Sebastes mentella | No directed fishery | all year | 13860 | TR | Pelagic in the Norwegian Sea, and as bycatch on the deep shelf and the continental slope | C, SG, MB |
| Sebastes marinus | GN, LL, HL | all year | 6300 | TR | Norwegian coast and southwestern Barents Sea | SG, MB MCS, MBU, C |
| Shrimp | TS | all year | 21053 |  | Svalbard, <br> Barents Sea, Coastal north of $62^{\circ} \mathrm{N}$ | ED, EF, SG, C, MCS |

${ }^{\text {A Provisional figures }}$
${ }^{\text {B }} \mathrm{On}$ a research quota
C The total cod catch north of $62^{\circ} \mathrm{N}(480,814 \mathrm{t})$ is the sum of the NEA cod catch given in the table above $(464,171 \mathrm{t})$ and the total cod catches between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ for the whole year and between $67^{\circ} \mathrm{N}$ and $69^{\circ} \mathrm{N}$ for the second half of the year $(16,643 \mathrm{t})$.
D The directed fishery for wolffish is mainly in ICES area IIb and the Russian EEZ, and the regulations are mainly restricted to this fishery
${ }^{\text {e }}$ Norwegian and Russian landings
${ }^{\mathrm{F}}$ The only directed fishery for Greenland halibut is by a limited Norwegian fleet, comprising vessels less than 28 m .

Table 1.11. Flexibility in coupling between the fisheries. Fleets and impact on the other species ( $\mathrm{H}, \mathrm{high}, \mathrm{M}$, medium, L, low and 0 , nothing). The table below the diagonal indicates what gears couples the species, and the strength of the coupling is given above the diagonal. The gears are abbreviated as: trawl roundfish (TR), trawl shrimp (TS), longline (LL), gillnet (GN), handline (HL), purse seine (PS), Danish seine (DS) and trawl pelagic (TP).

| Species | Cod | Coastal cod | Haddock | Saithe | Wolffish | S. mentella | S. marinus | Greenland halibut | Capelin | Shrimp |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Cod |  | H | H | H | M | M | M | M | L | M-H <br> juvenile cod |
| Coastal cod | TR, PS, GN, <br> LL, HL, DS |  | H | H | L | L | M-L | L | 0-L | L |
| Haddock | TR, PS, GN, LL, HL, DS | TR, PS, GN,LL, HL, DS |  | H | M | M | M | L | 0-L | M-H juvenile haddock |
| Saithe | TR, PS, GN, <br> LL, HL, DS | TR, PS, <br> GN,LL <br> HL, DS | TR, PS, GN, <br> LL, HL, DS |  | L | L | M | 0 | 0 | 0 |
| Wolffish | $\begin{aligned} & \text { TR, GN, LL, } \\ & \mathrm{HI} \end{aligned}$ | $\begin{aligned} & \hline \text { TR,GN, } \\ & \text { LL, HL } \end{aligned}$ | $\begin{aligned} & \text { TR, GN, LL, } \\ & \text { HL } \end{aligned}$ | $\begin{aligned} & \text { TR, GN, LL, } \\ & \text { HL } \end{aligned}$ |  | M | M | M | 0 | M juvenile wolffish |
| S. mentella | TR | TR | TR | TR | TR |  | M | H | H <br> juvenile Sebastes | H <br> juvenile <br> Sebastes |
| S. marinus | TR,GN, LL | TR,GN, LL | TR, GN, LL | TR,GN | TR, LL | TR |  | L | 0 | L-M juvenile Sebastes |
| Greenland halibut | TR, GN, LL,DS | TR,GN, LL | TR, GN, LL,DS | TR, GN, <br> LL,DS | TR, LL | TR | TR |  | 0 | M-H juvenile |
| Capelin | TR, PS, TS, TP | PS, TP | $\begin{aligned} & \text { TR, PS, TS, } \\ & \text { TP } \end{aligned}$ | PS | TP | TP | TP | None |  | L |
| Shrimp | TS | TS | TS | TS | TS | TS | TS | TS | TS |  |

Table 1.12. The averaged annual relative changes (\%) of cod fishable stock biomass under various combinations of cod stock size, capelin stock size and water temperature according to STOCOBAR long-term simulations (the harvesting strategy is based on Fpa=0,5, Bpa=460 thousand tonnes). Different colours denotes different natural potential of cod to stock changes: red is high potential to stock decline, yellow is low potential to stock change, and red is high potential to stock increase.

| Temperature$, \mathrm{C}^{\circ}$ | Cod FSB* averaged for 3 previous years, millions t | Capelin stock biomass, millions t |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | <1 | 1-3 | 3-5 | >5 |
| <3,6 $\mathrm{C}^{\circ}$ | <1,4 | -8,96 | -1,73 | 4,90 | 5,60 |
|  | 1,4-1,8 | -11,60 | -7,28 | -3,89 | 2,98 |
|  | >1,8 | -16,89 | -12,56 | -7,94 | -3,61 |
| 3,6-4,2 $\mathrm{C}^{\circ}$ | <1,4 | 1,17 | 5,14 | 12,82 | 15,99 |
|  | 1,4-1,8 | -7,29 | -3,96 | 2,91 | 6,72 |
|  | >1,8 | -12,24 | -8,52 | -5,24 | -0,27 |
| $>4,2 \mathrm{C}^{\circ}$ | <1,4 | 3,77 | 7,14 | 16,78 | 20,94 |
|  | 1,4-1,8 | -2,53 | 1,36 | 8,34 | 16,96 |
|  | >1,8 | -3,62 | -0,95 | 1,25 | 1,31 |

*Fishable stock biomass

Table 1.13. Overview of available prognoses of NEA cod recruitment (in million individuals of age 3) from different models (sections 1.4.5) together with the 2009 assessment estimates (ICES AFWG 2009 Table 1.12).

| Model | Prognostic years | Updated | 2010 <br> Prognoses | 2011 <br> Prognoses | 2012 <br> Prognoses | 2013 <br> prognoses |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Titov0 | 0 | At assessment | 480 |  |  |  |
| Titov1 | $1\left(2^{1}\right)$ | At assessment | 518* | 470 |  |  |
| Titov2 | 2 | At assessment | 451 | 323* |  |  |
| Titov3 | 3 | At assessment | 250* | 276* | 484* |  |
| Titov4 | 4 | At assessment | 425 | 362 | 780 | 946 |
| $\begin{aligned} & \text { TB (1984- } \\ & 2000) \end{aligned}$ | 3 | Last year assessment | 632 | 553 |  |  |
| $\begin{aligned} & \text { TB (1984- } \\ & 2004) \end{aligned}$ | 3 | Last year assessment | 627 | 551 |  |  |
| JES1 | $2\left(3^{2}\right)$ | At assessment | 878 | 797* | 827 |  |
| JES2 | $1\left(2^{2}\right)$ | At assessment | 714 | 669 |  |  |
| JES3 | $0\left(1{ }^{2}\right)$ | At assessment | 568 |  |  |  |
| H1 | 2 | At assessment | 890 | 889 |  |  |
| H2 | 2 | At assessment | 566 | 636 |  |  |
| H3 | 1 | At assessment | 500 |  |  |  |
| H4 | 1 | At assessment | 475 |  |  |  |
| $\begin{aligned} & \text { RCT3 } \\ & 2010 \end{aligned}$ | 3 | At assessment | 289 | 558 | 675 |  |
| Hybrid <br> Model <br> (Assessment 2009) |  | Last year assessment | 487 | 184 |  |  |
| Hybrid <br> model <br> (Assessment <br> 2010) |  | At assessment | 384 | 465 | 484 |  |

${ }^{1}$ Based on calculation of data from 2010.
${ }^{2}$ Based on prognosis estimate of capelin maturing biomass for October 1 2010, thereby allowing for an additional year.

* Models that are used in the Hybrid model at the 2010 assessment

Table 1.14. Proportion of cod in the diet of cod, based on Norwegian consumption calculations.

| Cod (predator)age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | $11+$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year |  |  |  |  |  |  |  |  |  |  |  |
| 1984 | 0.0000 | 0.0000 | 0.0032 | 0.0000 | 0.0437 | 0.0263 | 0.0328 | 0.0359 | 0.0367 | 0.0390 | 0.0374 |
| 1985 | 0.0015 | 0.0009 | 0.0014 | 0.0017 | 0.0314 | 0.0076 | 0.0827 | 0.0834 | 0.0842 | 0.0847 | 0.0853 |
| 1986 | 0.0000 | 0.0022 | 0.0015 | 0.0004 | 0.0130 | 0.1761 | 0.1767 | 0.1766 | 0.1762 | 0.1757 | 0.1748 |
| 1987 | 0.0000 | 0.0000 | 0.0007 | 0.0051 | 0.0103 | 0.0246 | 0.0377 | 0.0400 | 0.0418 | 0.0405 | 0.0436 |
| 1988 | 0.0000 | 0.0000 | 0.0000 | 0.0002 | 0.0058 | 0.0014 | 0.0038 | 0.0036 | 0.0032 | 0.0038 | 0.0036 |
| 1989 | 0.0000 | 0.0006 | 0.0016 | 0.0019 | 0.0027 | 0.0040 | 0.0035 | 0.0035 | 0.0040 | 0.0038 | 0.0041 |
| 1990 | 0.0000 | 0.0000 | 0.0000 | 0.0012 | 0.0017 | 0.0019 | 0.0268 | 0.0268 | 0.0268 | 0.0268 | 0.0268 |
| 1991 | 0.0000 | 0.0005 | 0.0000 | 0.0003 | 0.0032 | 0.0020 | 0.0224 | 0.0232 | 0.0235 | 0.0239 | 0.0241 |
| 1992 | 0.0000 | 0.0021 | 0.0037 | 0.0129 | 0.0250 | 0.0475 | 0.0120 | 0.0159 | 0.0232 | 0.0232 | 0.0230 |
| 1993 | 0.0000 | 0.0413 | 0.0368 | 0.0515 | 0.0536 | 0.1156 | 0.0498 | 0.0801 | 0.0801 | 0.0801 | 0.0805 |
| 1994 | 0.0000 | 0.0038 | 0.0917 | 0.0347 | 0.0285 | 0.0784 | 0.1247 | 0.1339 | 0.2617 | 0.2634 | 0.2608 |
| 1995 | 0.0069 | 0.0811 | 0.0745 | 0.0802 | 0.0925 | 0.1123 | 0.1389 | 0.2533 | 0.2553 | 0.2561 | 0.2574 |
| 1996 | 0.0000 | 0.1493 | 0.2549 | 0.2060 | 0.1322 | 0.1267 | 0.1850 | 0.2082 | 0.2459 | 0.2471 | 0.2465 |
| 1997 | 0.0000 | 0.0704 | 0.0767 | 0.1140 | 0.1552 | 0.1554 | 0.2329 | 0.2267 | 0.2882 | 0.2815 | 0.2832 |
| 1998 | 0.0000 | 0.0135 | 0.0272 | 0.0418 | 0.1041 | 0.0981 | 0.1081 | 0.1492 | 0.2758 | 0.2767 | 0.2778 |
| 1999 | 0.0000 | 0.0000 | 0.0049 | 0.0137 | 0.0148 | 0.0338 | 0.0620 | 0.1117 | 0.1937 | 0.1940 | 0.1840 |
| 2000 | 0.0000 | 0.0000 | 0.0286 | 0.0147 | 0.0134 | 0.0266 | 0.0499 | 0.0566 | 0.2757 | 0.2726 | 0.2738 |
| 2001 | 0.0000 | 0.0158 | 0.0116 | 0.0082 | 0.0131 | 0.0241 | 0.0496 | 0.0381 | 0.3294 | 0.3264 | 0.3301 |
| 2002 | 0.0000 | 0.0387 | 0.0591 | 0.0142 | 0.0187 | 0.0285 | 0.0359 | 0.0627 | 0.1603 | 0.1575 | 0.1581 |
| 2003 | 0.0000 | 0.0193 | 0.0198 | 0.0199 | 0.0206 | 0.0188 | 0.0456 | 0.1043 | 0.2257 | 0.2281 | 0.2269 |
| 2004 | 0.0230 | 0.0223 | 0.0294 | 0.0214 | 0.0184 | 0.0294 | 0.0391 | 0.0710 | 0.1059 | 0.1056 | 0.1061 |
| 2005 | 0.0000 | 0.0261 | 0.0229 | 0.0258 | 0.0155 | 0.0241 | 0.0487 | 0.0830 | 0.1688 | 0.1667 | 0.1693 |
| 2006 | 0.0000 | 0.0051 | 0.0007 | 0.0130 | 0.0285 | 0.0124 | 0.0397 | 0.0316 | 0.0841 | 0.0845 | 0.0834 |
| 2007 | 0.0000 | 0.0000 | 0.0010 | 0.0108 | 0.0137 | 0.0314 | 0.0336 | 0.0724 | 0.1518 | 0.1543 | 0.1504 |
| 2008 | 0.0000 | 0.0821 | 0.0243 | 0.0068 | 0.0089 | 0.0110 | 0.0820 | 0.1004 | 0.1223 | 0.1212 | 0.1198 |
| 2009 | 0.0376 | 0.0353 | 0.0227 | 0.0137 | 0.0147 | 0.0250 | 0.0981 | 0.0918 | 0.0920 | 0.0919 |  |
| Average | 0.0278 | 0.0339 | 0.0474 | 0.0673 | 0.0881 | 0.1437 | 0.1434 | 0.1432 |  |  |  |

Table 1.15. Cannibalism mortality in cod.

| Year | M at age 3 | M at age 4 |
| :---: | :---: | :---: |
|  | by regression |  |
| 2009 | 0.40 | 0.27 |
| 2010 | 0.43 | 0.28 |
| 2011 | 0.46 | 0.29 |
| 2012 | 0.64 | 0.36 |
|  | values used in assessment |  |
| 2010-2012 | 0.3335 | 0.227 |

Table 1.16. Overview of the standard sections monitored by IMR and PINRO in the Barents Sea, with observed parameters. Parameters are: T-temperature, S-Salinity, N-nutrients, chlachlorophyll, zoo-zooplankton, O-oxygen.

| Section | Institution | Time Period | ObSERVATION <br> FREQUENCY | PARAMETERS |
| :--- | :--- | :--- | :--- | :--- |
| Fugløya-Bear <br> Island | IMR | 1977-present | 6 times pr year | T,S,N,chla,zoo |
| North cape-Bear <br> Island | PINRO | 1950's-present | yearly | T,S |
| Bear Island-East | PINRO | 1950's-present | yearly | T,S |
| Vardø-North | IMR | 1977-present | 4 times pr year | T,S,N,chla |
| Kola | PINRO | 1921-present | monthly | T,S,O,N |
| Kanin | PINRO | 1950's-present | yearly | T,S |
| Sem Islands | IMR | 1970's-present | Intermittently* | T,S |

* The Sem Island section is not observed each year, and have not been observed the last 3-4 years.

Table 1.17. Overview of conducted monitoring surveys by IMR and PINRO in the Barents Sea, with observed parameters and species. For zooplankton, mammals and benthos abundance and distribution for many species are investigated. Therefore, in the table it is only indicated whether sampling is conducted. Climate and phytoplankton parameters are: T-temperature, S-Salinity, N-nutrients, chla-chlorophyll.

| Survey | InSTITUTION | Period | Climate | PhytoPLANKTON | ZOO-PLANKTON | JUVENILE FISH | TARGET FISH STOCKS | Mammals | Benthos |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Winter | Joint | Feb-Mar | T,S | N, chla | intermittent | All commercial species and some additional | Cod, Haddock | - | - |
| Lofoten | IMR | Mar-Apr | T,S | - | - |  | Cod, haddock, saithe | - | - |
| Ecosystem survey | Joint | Aug-Oct | T,S | N,chla | Yes | All commercial species and some additional | All commercial species and some additional | Yes | Yes |
| Norwegian coastal surveys | IMR | Oct-Nov | T,S | N,chla | Yes | Herring, sprat, demersal species | Saithe, coastal cod | - | - |
| Autumnwinter trawlacoustic survey | PINRO | Oct-Des | T, S | - | Yes | Demersal species | Demersial species | - | - |
| Norwegian Greenland halibut survey | IMR | Aug | - | - | - | - | Greenland halibut, redfish | - | - |
| Russian young herring survey | PINRO | May | T, S |  | Yes |  | Herring | - | - |



Figure 1.1. The main features of the circulation and bathymetry of the Barents Sea.


Figure 1.2. Temperature (upper) and salinity (lower) anomalies in the $50-200 \mathrm{~m}$ layer of the Fugløya-Bear Island section (left) and the Vardø-North section (right).


Figure 1.3. Monthly temperature anomalies in the $\mathbf{0 - 2 0 0} \mathrm{m}$ layer of the Kola Section in 2008 (black dots) and 2009 (red bars). St. 1-3 - Coastal waters, St. 3-7 - Murman Current (Anon., 2010).


Figure 1.4. Bottom temperature anomalies in the Barents Sea in August-September 2009 (Anon., 2010).


Figure 1.5. Observed Atlantic Water volume flux through the Fugløya-Bear Island section estimated from current meter moorings. Three months (blue line) and 12-months (red line) running means are shown.


Figure 1.6. Ice extent anomalies in the Barents Sea in 1982-2009 (Anon., 2010). The blue line shows monthly values, the red one -11 -month moving average values.


Figure 1.7. Horizontal distribution of zooplankton in 2009 ( $\mathrm{g} \mathrm{m}^{-2}$ of dry weight from bottom-0 m).


Figure 1.8. The mean abundance of euphausiids in the north-western and western areas of the Barents Sea in 2008 and 2009.


Figure 1.9. Biomass of pelagic fish species in the Barents Sea. Data are taken from; capelin: Acoustic estimates in September, age 1+ (ICES AFWG 2010), herring: VPA estimates of age 1 and 2 herring (ICES C.M. 2009/ACOM:12), using standard weights at age ( 9 g for age 1 and 20 g for age 2); polar cod and blue whiting: Acoustic estimates in September, age 1+ (Anon. 2010), 0-group: estimates of biomass of cod, haddock, herring and capelin 0-group, corrected for catching efficiency (Eriksen et al. 2010).


Figure 1.10. Biomass of demersal fish species in the Barents Sea. Data are taken from; cod: VPA estimates, age 3+ (ICES AFWG 2010); haddock: VPA estimates, age 3+ (ICES, 2010); Greenland halibut: VPA estimates, age 5+ (ICES, 2007); Sebastes mentella: VPA estimates, age 6+ (ICES, 1995 for the years 1968-1990; ICES, 2003 for the years 1991-2002).


Figure 1.11a. Time series of annual average fishing mortalities for Northeast Arctic cod (time period 1946-2009, average for ages 5-10), Northeast Arctic saithe (time period 1960-2009, average for ages 4-7), coastal cod (1984-2009, average for ages 4-7), Northeast Arctic haddock (time period 1950-2009, average for ages 4-7), Greenland halibut (time period 1964-2009, average for ages 6-10) and Sebastes marinus (time period 1990-2009, average for ages 12-19).


Figure 1.11b. Left panel - annual fishing mortalities of the Northeast Arctic cod, haddock and saithe stocks relative to the critical levels above which the fishing mortality will impair the recruitment. Right panel - annual fishing mortalities of Golden redfish (Sebastes marinus) and Greenland halibut (Reinhardtius hippoglossoides) relative to the proposed maximum levels above which the fishing mortality over time most probably will impair the recruitment.


Figure 1.12. Pair-wise plots of annual average fishing mortalities (above diagonal) and landings (below diagonal) for overlapping time periods for Northeast Arctic cod (time period 1946-2009, average for ages 5-10), Northeast Arctic haddock (time period 1950-2009, average for ages 4-7), Northeast Arctic saithe (time period 1960-2009, average for ages 4-7), coastal cod (1984-2009, average for ages 4-7), Greenland halibut (time period 1964-2009, average for ages 6-10) and Sebastes marinus (time period 1990-2009, average for ages 12-19). The correlation and the corresponding pvalue are given in the legend.


Figure 1.13. Relative distribution by weight of cod, haddock, saithe, Greenland halibut, golden redfish (Sebastes marinus), beaked redfish (Sebastes mentella) and other species taken by Russian bottom trawl in 2009 per main area for the Russian strata system.


Figure 1.14. Relative distribution by weight of Norwegian catches of cod, haddock, and saithe per main area for the Norwegian strata system.


Figure 1.15. The Russian catch of cod, haddock, saithe, Greenland halibut, Sebastes marinus, Sebastes mentella and other species taken by bottom trawl by main statistical areas in 2009, thousand tonnes. The statistical areas correspond to the areas shown in Figure 1.13.


Figure 1.16. The Norwegian catch of cod, haddock and saithe by main statistical areas in 2009, thousand tonnes. The statistical areas correspond to the areas shown in Figure 1.14.


Figure 1.17. Upper panel - gear composition of the Norwegian groundfish (2007; left) and pelagic capelin (2000-2008; right) fisheries in the Northeast Arctic. Note that the purse seine in the groundfish fishery is solely used in a coastal fishery for saithe. Lower panel - gear composition of the Russian groundfish (2007; left) and pelagic capelin (2000-2008; right) fisheries in the Northeast Arctic.


Figure 1.18. Positions of the standard sections monitored in the Barents Sea. A is fixed station Ingøy, B is Fugløya-Bear Island, C is North cape-Bear Island, D is Vardø-North, E is Kola, F is Sem Island-North $G$ is Kanin section and $H$ is Bear Island-East section.

## 2 Cod in subareas I and II (Norwegian coastal waters)

Type of assessment: The schedule says "Update", but several revisions are presented. The main reason for the revisions has been to have some additional basis for evaluating the proposed rebuilding plan.

A new catch series for recreational and tourist fisheries is presented and added to the commercial catch. The combined data series is found to fit better with the survey, using stock dependent catchability for ages 2 and 3 in a XSA otherwise set to standard values given in the Quality Handbook Stock Annex. General information regarding the stock and earlier assessments are given in an updated Quality Handbook Stock Annex.

A rebuilding plan has been proposed by Norwegian authorities and has been evaluated by the working group.

### 2.1 Fisheries

Coastal cod is to a variable extent fished throughout the year and within nearly all the distribution area (inside the 12 n.mile zone in the Norwegian statistical areas 03, $04,05,00,06,07$, Figures 2.1-2.3). The main fishery for coastal cod takes place in the first half of the year. The main fishing areas are along the coast from Varangerfjord to Lofoten (areas 03, 04, 05, 00).

Except for the open fjords in eastern Finnmark, the quantities fished inside fjords are quite low. The total share between gear types in the estimated coastal cod commercial landings has in recent years been around $50 \%$ for gillnet, $20 \%$ for Danish seine, 20\% for long-line/hand-line and less than $5 \%$ for bottom trawl.

Recreational fisheries take an important fraction of the catches in some local areas, especially near the coastal cities and in some fjords where commercial fishing activity is low. There is no measurements of the amount of Norwegian coastal cod (NCC) taken by recreational or tourist fishers in Norway. However, there are a few reports trying to assess the amount in certain years and these reports have been used to construct time series based on assumptions made in the reports of temporal trends.

A survey for mapping recreational fisheries was conducted in 2003 (Hallenstvedt and Wulff, 2004) and the results from this report gives reason to assume that there were fished app. 13,000 $t$ of cod by recreational fishers in 2003 north of $62^{\circ} \mathrm{N}$. This is based on $50 \%$ of the catches in the area being cod and that due to the fishing season almost all of the cod is coastal cod. Nedreaas (2005) discuss this assumption and assumes that the winter fishery by recreational fishers only consists of North east arctic cod. This is probably not the case - since the winter fishery is small and is probably conducted close to home.

The effort used in recreational fisheries is monitored through surveys of questionnaires mapping the amount of the population that has conducted recreational fisheries during the last year and to what extent they have fished in salt water or freshwater. Based on interpolating these surveys onto the development of the population in Norway, it is possible to give an index of effort in recreational fisheries in the sea. It is assumes that recreational fisheries are conducted to catch a desired amount of fish - and that the effort is not restricted in time. This gives the quantity taken to be proportionate to the effort - and not influenced by the stock size.

Some recreational fishers deliver their catches to the sales organisations. In this working document it is assumed that this group is not included in the interview material and that these landings are already included in the reported catches from the commercial fisheries. This is also contradictory to the conclusions from Nedreaas (2005).

Thus, the quantity of $13,000 \mathrm{t}$ NCC is assumed to be taken by the recreational fishers in Norway in 2003 has been extrapolated to the years before and after using the product of population numbers and the fraction of the people conducting recreational sea fisheries. It is assumed that the amount of cod is $50 \%$ throughout all the years.

There is one report available to indicate the level of tourist fisheries in Norway. The report is by the consultant company Essens management (Anon, 2005) and is based partly on Hallenstvedt and Wulff, 2004 and partly by surveys on the number of tourists who say they have been fishing in the sea.

This report estimates the tourist fishery north of $62^{\circ} \mathrm{N}$ for cod to amount to $1,100 \mathrm{t}$ in 2004. They also assume that the increase in tourism for sea fishing increased with $19 \%$ per year from 1995 until 2000, then increased with $16 \%$ per year until 2004. In this working document it is assumed that the increase until 2009 has been $10 \%$ per year. This gives a quantity in 2009 of $1,800 \mathrm{t}$ cod. It also gives a time series back to the beginning of the 1990s assuming that the catch is proportional to the number of tourists fishing in the sea.

There are ongoing investigations of tourist fisheries and the results of these investigations will only be available at a later time. However, there is reason to believe that the figure of 1800 t cod is not out of scale with the ongoing investigations (pers. comm. Nedreaas, 2010).

The constructed time series may not be as accurate as desired, however, the level of catch to be added to the commercial catches is assumed to be fairly well documented. Also the trend in both the recreational fisheries and tourist fisheries seem to be consistent with what has been presented in later years.

### 2.1.1 Sampling fisheries and estimating catches (Tables 2.1-2.2)

The commercial catches of Norwegian Coastal cod (NCC) have been calculated back to 1984 (Table 2.1a). For this period the estimated landings have been between 22,000 and $75,000 \mathrm{t}$. The estimated landings of NCC in 2008 are $25,777 \mathrm{t}$ and in 2009 they are $24,821 \mathrm{t}$ (Table 2.1a, Figure 2.4). Table 2.1 b shows the estimated catch by gears, area and quarters in 2009.

In table 2.1c is shown the age distribution for long line and hand line raised to the combined estimates of recreational and tourist fisheries, together with the two estimated time series for these two fisheries.

Commercial catches of cod are separated to types of cod by the structure of the otoliths in commercial samples. Figure 2.5 illustrates the main difference between the two types: The figure and the following text is from (Berg et al., 2005): Coastal cod has a smaller and more circular first translucent zone than north-east Arctic cod, and the distance between the first and the second translucent zone is larger (Fig. 2.5). The shape of the first translucent zone in north-east Arctic cod is similar to the outer edge of the broken otolith and to the subsequent established translucent zones. This pattern is established at an age of 2 years, and error in differentiating between the two major types does not increase with age since the established growth zones do not change with age. The precision and accuracy of the separation method has been investigated by comparison of different otolith readers and results from genetic investigation of cod. The results indicate high accuracy
using in the otolith method (Berg et al., 2005). Nevertheless, in cases with a low percentage misclassification of large catches of pure NEA cod, the catches of coastal cod could be severely overestimated.

The basis for estimating coastal cod catches is the total landings of cod inside the 12 n . mile zone in the Norwegian statistical areas 03, 04, 05, 00, 06, 07 (Figures 2.1-2.3), combined with the sampling of these fisheries. Tables 2.2 and 2.3 show the sampling of the cod fishery by quarters and areas in 2009 and earlier. The total number of age samples was 359 . Since the catches are separated to type of cod by the structure of the otoliths, the numbers of age samples are critical for the estimated catch of coastal cod. A total of about 11,000 fish were aged. More than 2,600 of these otoliths were classified as coastal cod.

Table 2.4 shows the estimated catches of coastal cod by statistical area and quarter for the years 2006-2009. The corresponding fractions of coastal cod in cod catches are also shown. In the southern areas $(06 / 07)$ the proportions are close to 1.0 in all quarters, except for some years when some NEA cod spawn far to the south in quarter 1 and 2. In the other areas the proportions are lower in quarter 1 and 2 in all years due to the spawning migration of NEA cod. In area 03 (eastern Finnmark) a considerable proportion of NEA cod is present also during autumn.

The calculation of coastal cod landings for recent years has been problematic for parts of the Lofoten area. This relates to the Norwegian statistical area 00 (outer Vestfjord, the area south of Lofoten archipelago, Figure 2.3) in quarter 1 and 2. This area has historically been an important spawning area for Northeast Arctic cod. In the period 2004-2009 a major part of the Northeast Arctic cod was spawning in the outer, southwestern part of the area, and almost nothing in the north-eastern part. Most of the commercial catches in the area were taken in the south-western part (locations 03 and 04, Figure 2.3) where the density of cod was much higher than in the north-eastern part. In the same period the sampling intensity has been highest for the catches in the north-eastern part (locations 46 and 48) where coastal cod dominated. (In most of this north-eastern area the fishery was restricted to vessels below 15 m and use of Danish seine was not allowed). The catch sampling has not been sufficiently accurate to split the catches between those locations. Merging all samples in the whole area is therefore considered to overestimate landings of coastal cod. In order to obtain a more realistic catch in the area for the years 2004-2009, the working group in the years from 2007 has used only the samples taken from the south-western part for separating the total catch in the area between coastal cod and Northeast Arctic cod. The recorded positions of the samples are considered to be accurate.

### 2.1.2 Regulations

The Norwegian cod TAC is a combined TAC for both NEAC stock and NCC stock. The coastal cod part of this combined quota was set 40,000t in 2003 and earlier years. In 2004 it was set to $20,000 t$ t, and in the following years to $21,000 \mathrm{t}$. There are no separate quotas given for the coastal cod for the different groups of the fishing fleet.

Trawl fishing for cod is not allowed inside the 6-n.mile. Since the mid 1990 the fjords in Finnmark and northern Troms (areas 03 and 04) has been closed for fishing with Danish seine. Since 2000 the large longliners have been restricted to fish outside the 4 n.mile.

For the fisheries in 2010 there were also set a quota for recreational and tourist fisheries together with the quota for young fishers, to be $10,000 \mathrm{t}$ and allocated from the agreed TAC.

To achieve a reduction in landings of coastal cod additional technical regulations in coastal areas were introduced in May 2004 (after the main fishing season) and continued with small modifications in 2005 and 2006. In the new regulations "fjord-lines" are drawn along the coast to close the fjords for direct cod fishing with vessels larger than 15 meter. Further restrictions were introduced in 2007, 2008 and continued in 2009, by not allowing pelagic gill net fishing for cod and by reducing the allowed bycatch of cod when fishing for other species inside fjord lines from $25 \%$ to $5 \%$, and outside fjord-lines from $25 \%$ to $20 \%$. A box closed for all fishing gears except handline and fishing rod has since 2005 been defined in the Henningsvær-Svolvær area. A similar box has in 2009 and 2010 been closed during the spawning season in Borgundfjord near Alesund. These are areas where spawning concentrations of coastal cod is usually observed and where the catches of coastal cod has been high. Since the coastal cod is fished under a merged coastal cod/north-east arctic cod quota, these regulations are supposed to turn parts of the traditional coastal fishery over from catching coastal cod in the fjords to catch more cod outside the fjords where the proportion of Northeast Arctic cod is higher.

### 2.2 Survey data

A trawl-acoustic survey along the Norwegian coast from the Russian boarder to $62^{\circ} \mathrm{N}$ was started in the autumn 1995. In 2003 the survey was somewhat modified by being combined with the former saithe survey at the coastal banks and the survey was moved from September to October-November. This new survey covers a larger area than the coastal surveys in 1995-2002. However, the survey indices for cod to be used in this report are calculated using the same area coverage and the same method as in the years previous to 2003.

### 2.2.1 Indices of abundance and biomass (Tables 2.5-2.11, Figs 2.7 to 2.13)

The results of the 2009 survey (Mehl et al. WD 8 2010) are presented in Tables 2.5-2.11 for the area inside the 12 n . miles border in the Norwegian statistical areas $03,04,05$, 00, 06, and 07 (Figures 2.1 and 2.2). The survey time series of estimated numbers of NCC per age groups is given in Table 2.6. For most age groups the estimates are close to the lowest ever observed, and the total number is the third lowest observed. The 2009 estimate of survey biomass is about 39,100 t (Table 2.9) and this is an increase from last year, and is mainly due to an increase in most age groups except 2 and 3 year olds. The estimated spawning biomass is $13,500 \mathrm{t}$, and this is the lowest observed (Tables 2.11). However, this is mainly due to a downward change in the maturity at age, which may have been influenced by an earlier start of the survey this year, by two weeks. The bulk of the spawning biomass is comprised of ages 5-7.

Similar changes in maturity have been observed at earlier surveys. Variable timing of the survey and annual variation of the onset of the maturation process could be parts of the reason for this. Since the rebuilding plan is made conditional to the survey estimate of SSB, it would be more robust to use a more smoothed, or even fixed ogive for calculating the survey SSB. This approach will give an increased biomass from 2008 to 2009.

The $4+$ biomass (summed from Table 2.9) is plotted together with total biomass and spawning biomass in Figure 2.13.

The pattern seen (Figure 2.6) over the full time series of abundance at age is that ages 2 and 3 have declined more, and over a longer period, compared to the older fish. The
series now indicates a rather stable stock at a low level. The period since 2002 shows considerable variation, however, without any trend.

Figures 2.7-2.12 show the time series of stock number within each statistical area. In areas 03,04 and 05 the decline since the late 1990s is rather parallel. In the other three areas the year-to- year variation is larger, but similar trends are indicated. These latter southern areas contribute less to the total estimate.

### 2.2.2 Age reading and stock separation (Tables 2.2, 2.3, 2.8-2.10)

A total of 2341 cod otoliths were sampled during the 2009 survey.
As in previous years, NCC was found throughout the survey area. The 2009 survey data on the stock separation are similar to the 2007 and 2008 data and shows the same pattern as the whole 1995-2008 surveys. The sampling showed a higher proportion of NCC in the fjords and to the south compared with the northern and outer areas. The proportion of the NCC increases going from north to south along the Norwegian coast. Table 2.12 show the proportions of coastal cod in the survey samples by age and statistical areas in 2008. Nearly all otoliths collected south of $67^{\circ} \mathrm{N}$ (Norwegian statistical areas 06 and 07) were NCC type. Although the proportions are lower, the total abundance of NCC is higher north of $67^{\circ} \mathrm{N}$ (Table 2.5).

Table 2.12 also show the proportions of coastal cod in the survey samples by age for 5 previous years. The proportion is rather stable between years, but is consistently higher for young fish compared to old.

It must be emphasised that the Norwegian coastal surveys is conducted in OctoberNovember, and there is usually more NEA cod in the coastal areas at other times of the year, especially during the spawning season in the late winter. This is reflected in the commercial sampling as shown in Table 2.4.

### 2.2.3 Weights at age (Table 2.8)

As observed in the earlier surveys, there is a general tendency for costal cod to have higher weight at age when caught in the southernmost area. Table 2.8 show the time series of mean weights at age for the whole survey.

### 2.2.4 Maturity-at-age (Table 2.10)

The maturity-at-age is estimated annual from the data collected at the coastal survey. The age at $50 \%$ maturity $\left(\mathrm{M}_{50}\right)$ for the NCC was near 5 in 2006, 2007 and 2008 surveys (Table 2.10), but increased to almost 7 years of age in 2009 . Both the estimated weights at age and the estimate of maturities are influenced by uncertain values in areas where few fish are sampled. In addition, the survey is conducted in the period October/November, a period when maturation stages are difficult to interpret. Therefore, much of the year to year variation observed might not be real, and a fixed long term average could be a reasonable alternative. In 2009 the survey was started two weeks earlier than the years before, and this may also have influenced the sampling of mature fish.

### 2.3 Data available for the Assessment

### 2.3.1 Catch at age (Table 2.1, and table 2.14)

The estimated commercial catch at age (2-10+) for the period 1984-2009 is given in Table 2.1a. Also, the estimated catches from recreational and tourist fisheries are given in table 2.1c. The combined catches as are given in table 2.14.
The total catches of coastal cod have been severely underestimated in earlier reports. In addition to the official landings from commercial vessels, an unknown amount of coastal cod is landed both from tourist fishing and from recreational fishing activity by Norwegian citizens. Two different investigations have estimated the amount of cod landed from these two activities and the reports were published in 2004 (in Norwegian). A summary of these two reports was presented as a WD to the 2005 WG (WD 23).

In this year's report, a new evaluation of the catches from recreational and tourist fisheries are given, based on the same reports but giving slightly different figures than in the 2005 WG report. The catch of coastal cod in 2003 is estimated to approximately 13,000 tonnes from the recreational fishing activity and in 2004 to be 1,100 tonnes from the tourist fishing. These figures sum up close to $50 \%$ of the official landings of coastal cod in 2004.

There have also been conducted two investigations trying to estimate the level of discarding and misreporting from the coastal vessels in two periods (2000 and 20022003, WD 14 at 2002 WG). The amount of the discard was calculated and the report from the 2000-investigation concluded there was both discard and misreport by species in 2000. Landings of cod with gillnet should be increased by approximately 8 $10 \% .1 / 3$ of this is probably Coastal cod. The last report concluded that misreporting in the Norwegian coastal gillnet fisheries have been reduced significantly since 2000.

From Hallenstvedt and Wulff (2004) it is seen that in the northern part of Norway almost no gill net fishing is included in the recreational fisheries. It is therefore reasonable to use the samples from long line and hand line to split the catches into age. The available material for coastal cod is for the whole year and this is used to split the estimated figures in tonnes to numbers at age.
For the early part of the time series for long line and hand line there are a large portion of the samples being aged 10 year and older. It is assumed that this is mainly from the winter fisheries for cod and therefore the $10+$ group is excluded from the material used to raise the numbers at age to the recreational and tourist catches.. This is also supported by a fairly low numbers of 9 year olds in that part of the material. In view of this it is assumed that it would be reasonable to assume that most of the recreational fishery is for fish younger than 10 year of age.

It seems to be clear that the commercial catches using hook and line reflect a severe failure in recruitment during the time series and anecdotal information seem to support this also for the recreational fishery. Recreational fishers frequently say that fishing grounds are no longer giving any yield, and that smaller cod is not available to fishers using fishing rod from land.

This matrix of recreational and tourist catches is proposed as a first solution to the problem that the commercial catches do not reflect the total amount being caught.

### 2.3.2 Weights at age (Tables 2.8 and 2.13)

Weight at age in catches is derived from the commercial sampling and is shown in Table 2.13. The same weight at age is assumed for the recreational and tourist catches.

The weight-at-age in the stock is obtained from the Norwegian coastal survey (Table 2.8). The survey is covering the distribution area of the stock. Weight-at-age from the survey is therefore assumed to be a relevant measure of the weight-at-age in the stock at survey time (October). These weights will, however, overestimate the stock biomass at start of the year (Table 2.13).

### 2.3.3 Natural mortality

A fixed natural mortality of 0.2 has been assumed in the assessment. However, in the Barents Sea cod cannibalism has been documented to be a significant source of mortality that varies in relation to alternative food and in relation to the abundance of large cod. This might also be the case for the coastal cod (Pedersen and Pope, 2003a and b). In the 2005 coastal cod survey 1125 cod stomachs were analysed (Mortensen 2007). The observed average frequency of occurrence of cod in cod stomachs was around $4 \%$. Other important predators on cod in coastal waters are cormorants and otters (Pedersen et al., 2007). Young saithe (ages 2-4) has been observed to consume postlarvae and 0-group cod during summer/autumn.

### 2.3.4 Maturity-at-age (Tables 2.10, 2.13)

The average maturity at age observed over the whole survey period (1995-2009) has been used in the assessment (Table 2.13).

### 2.4 Methods used for assessing stock trends

Earlier attempts to assess the stock using XSA analysis have not given reliable results. In the two last years the main basis for assessing the stock is the survey time series plotted in Figures 2.6-2.13, and a SURBA was used for further analysing the survey trends. However, this method is not recommended and did not give reliable results.

In this year's WG with the updated catch figures for recreational and tourist catches an attempt to use the XSA method was again tried. Four set of runs were performed. First, last year's data was updated with only the commercial catches and an XSA with the standard settings from the stock annex was run. Secondly, the new time series of catch on numbers and catch in tonnes were applied and the settings were as in the first run. Thirdly, an attempt to use stock dependant catchability for the two youngest ages (age 2 and 3) were done, and this XSA run was then taken to be the basic run for the assessment.

An additional analysis of trends in fishing mortality is presented in Annex 9. Mortality from $\log$ catch ratios in the canum matrix was calculated and survey mortalities were calculated from the coastal survey. These mortalities were regressed with the xsa-Fs and thereby used for estimating Fs.

The result of these evaluations was that F had remained fairly stable over the last 5 years, and an average value just above 0.35 was indicated for the analysis relating to commercial catch. The analysis relating to the new total catch data show a stable or slightly declining trend with F2009 close to 0.30 . On this basis it was decided to use the Fold from the third XSA together with the Fnew set as the average values of the years 2005 to 2008 in the same XSA for ages 3 and older to run a Standard VPA. For
age 2 the value from the XSA in 2009 was used. This fourth run, the SVPA, was then used as the final stock trends and further used in the predictions.

### 2.4.1 Input, diagnostics and results for the XSA tuning (Table 2.6, tables 2.13 -

 2.20)The data for the tuning is the survey indices given in Table 2.6. The diagnostic outputs from the three XSA runs are given in tables $2.15,2.17$ and 2.19. The summary tables are given in tables 2.16, 2.18 and 2.20. SSB-values refer to 1 January.

### 2.4.2 Input for the predictions (Tables 2.23 and 2.24)

Input to the predictions is set as the standard choices in the MFDP program. The fishing pattern is raised to F4-7 level in year 2009. Two sets of prediction are made, one with a fixed fishing pattern, and one with a fishing pattern in 2011 and onwards set with reduced F values on ages $2,3,4$ and 5 to accomplish a reduction in F4-7 of $15 \%$. This fishing pattern will allow less small fish to be caught. In order to give management options tables corresponding to the proposed rebuilding plan for Norwegian Coastal Cod, the factors for the intervals of F is set to vary by 0.15 , approximating a $15 \%$ change from one year to the next. The two input scenarios are given in tales 2.26 and 2.29 .

### 2.5 Results of the Assessment (Tables 2.21-2.22)

### 2.5.1 Comparing with last year's assessment (Figures 2.13 and 14)

Last year's assessment was based on the survey. In figure 2.13 is shown the development of total biomass, biomass of age 4 and older and the spawning biomass. Also figure 2.14 show the ratio of yield to the biomass of 4 year and older ion the survey.

In these figures are also added the SSB and F from this year assessment for comparison. The biomass development seem to be similar, however, the SSB in the SVPA is calculated with constant maturity ogive, whereas the maturity in the survey show a declining trend with time. The F4-7 from the VPA is more stable then the Yield/Biomass4+ in the survey. In addition to the survey variability, this may be due to an increase of ages 2,3 and 4 in the catches in the years 2002, 2003 and 2008, where high values for Yield/Biomass4+ is found.

### 2.5.2 Fishing mortalities and final Standard VPA (standard plots)

The fishing mortality (F4-7) shows a declining trend since 1999 and is now close to 0.30 . The VPA analysis reflects the increase in F4-7 in the years before 1999, which is also seen from the Yield/Biomass4+ in the survey.

From the retrospective plots of the XSA (Figure 2.15) one may see that there is a tendency for the F4-7 in the last year to be overestimated, and this is dealt with in the final SVPA by setting a terminal F based on external analysis (Annex 10). The terminal values used just happened to fit with the average values of the years 2005-2008 in the XSA.

### 2.5.3 Recruitment

The survey estimates of young age groups (1-3) in 2009 are among the lowest in the series, as were in particular the values in 2008. For ages 1, 2 and 3 the 2008 value was only about $1 / 10$ of the peak values in 1995, 1996 and 1997. At present there are therefore poor prospects for any rapid rebuilding of the stock in near future.

It is worth noting that the recruitment started to decline a few years before the spawning stock, indicating that the recruitment failure is the cause for the stock decline. Whether this recruitment is now at a stable level and will be maintained or if there is going to be a further decrease, is difficult to say. In the prediction is used a resent average of 21.5 millions (taken as the average of the 4 last years in the SVPA) and no further attempt to estimate the recruitment was made.

### 2.5.4 Catch options for 2011 and 2012 (Tables 2.25 and 2.26)

The results of the predictions are given in tables 2.25 and 2.26 . The second option where the fishing pattern is changed towards larger fish seems to give comparable results to the first option where the fishing pattern was kept constant. Both options give an increase of the SSB by reducing $\mathrm{F}_{4} 7$ by $15 \%$ each year over two years.

### 2.6 Comments to the Assessment

The acoustic survey probably has a larger relative uncertainty in later years compared to earlier. This is because cod now contributes to a lower fraction of the total observed acoustic values. The cod estimate is thus more vulnerable to allocation error. The Norwegian coastal survey is the only survey covering the distribution area of the stock. The survey is conducted in the period October/November. In this period the maturity can be difficult to define exactly and might influence the estimation of maturity-at-age and hence the estimation of SSB.

The new series with recreational and tourist fisheries included may be said to scale the stock to an more realistic level and give reason to believe that regulations according to this level may affect the stock in the desired direction.

The XSA tuning with stock dependant catchability on ages 2and 3 gave better fit to the survey data. The average survey biomass for the years 1995 to 1998 is defined as a preliminary target for a rebuilding plan.

### 2.7 Reference points

The analyses made for evaluating the Rebuilding Plan (Annex 10) also give some information regarding reference points. The assessment based on commercial catch plus recreational catch gives a stock-recruit break point at 139 kt SSB . The corresponding Fcrash is estimated to 0.38 .

The stock-recruit development may indicate that recruitment conditions may have changed. Assuming that increased SSB will not give recruitments higher than those observed for the year-classes 2000-2005, we get a break point at 103 kt . This is a reasonable candidate for Blim. The corresponding F crash is 0.32 , which is a candidate for Flim. F0.1 is estimated to 0.16. A safe long term Fmsy-target also has to be rather close to 0.16 . A corresponding MSY Btrigger would be in the range 150-200 kt. These MSY considerations are quite preliminary (see also section 0.10).

### 2.8 Management considerations

Catches have remained rather stable since 2004. The regulations seem to have reduced the catches compared to pre 2004 level but have not been sufficient to cause further reduction. The time series of recreational catch show rather stable catches, and they represent thereby a higher fraction (about 35\%) after 2004 compared to before.

### 2.9 Evaluation of Rebuilding plan for coastal cod

Annex 10 describes the analysis made for evaluation and presents the results of various simulations. The conclusions are;

If the plan is fully implemented it will lead to a safe rebuilding. Under presumed realistic errors a rather long rebuilding period is required, but the fishing mortality comes down to fairly safe levels within few years. On this basis the proposed rule is considered to be in accordance with the Precautionary Approach. Increasing the F step or aiming for annual reduction unconditional to survey results for the first 3-5 years will contribute to a faster and safer rebuilding. If future observations show recruitment declines stronger than assumed in the current stock-recruit model, the plan may need revisions. The new data on recreational fisheries also highlights the need to consider further regulation on these activities to obtain the F-reductions specified in the plan.

The current regulations aiming for protection of local stock components should be maintained. This should be improved when the scientific basis is improved.

Analyses were made both using input from the assessment based on only commercial catch and the one using all catches. For each of these data sets two recruitment scenarios were assumed; one with continued low recruitment near the recent average (2000-2005 year-classes), and one using the full historic recruitment series. In terms of rebuilding period needed to reach a safe F level, the analysis seemed fairly robust against choice of data sets and recruitment assumptions. The resulting stock size and catches was however higher in the cases where higher recruitment was assumed.

The main findings are that uncertainty in the survey and uncertainty in the implementation of F-reductions are both contributing to slow down reduction rate for the realized F and corresponding slow growth in stock. The general patterns were similar in all these simulation (see tables 4-9 in Annex 10). The series based on all catches and assuming recruitment restricted to recent levels were considered most relevant. With $15 \%$ reduction steps for $F$ the resulting time span from now (2010) was (Table 4 in Annex 10):
-about 7 years were needed to have high probability for F being below F crash
-about 10 years needed for average SSB above rebuilding target
-about 15 years to have high probability for $\mathrm{SSB}>\mathrm{B}_{\lim }$
Larger steps in the reduction rates or making the reduction every year (unconditional to the survey result) will speed up the process somewhat.
The unconditional case gives (Table 8 in Annex 10):
-about 4 years were needed to have high probability for F being below F crash
-about 10 years needed for average SSB above rebuilding target
-about 10 years to have high probability for $S S B>B_{l i m}$
The precautionary criteria of high probability of F below Fcrash and SSB above Blim seem achievable within a reasonable time frame. Provided that the management is able to enforce regulations that are efficient in reducing $F$, the plan is considered to be in accordance with the precautionary approach. Further actions will be needed if there are signs that recruitment declines further.

The new data on recreational fisheries also highlights the need for further regulation on these activities to follow the planned F-reduction.

In these quantitative simulations and analyses no direct attempts have been made to take account of the stock complexity. Genetic studies indicate that the cod in some fjords could be separate stocks isolated from neighboring stocks. An assessment of the merged stock is not likely to detect fluctuations of the smaller components, and thereby the current assessment approach involves some risk to local stocks. The stock complex is still not fully mapped, but the existence of local stocks also calls for special attention for protecting genetic diversity. Full monitoring and research on small local stocks requires large efforts and may not be realistic. A possible approach could be to obtain information from local fisheries and look for data that could be appropriate indicators for at least detecting sharp declines of local stocks. The established strategy of more strict regulations inside the fjords than outside should be continued.

A fixed natural mortality of 0.2 is used both in the assessment and the simulations. Some fjord studies (Pedersen and Pope, 2003a and b, Mortensen 2007, Pedersen et al., 2007, Aas, 2007) indicate that the main predators on young cod are larger cod, cormorants and saithe. There are no estimates of annual predation mortality for the stock complex. Thus, the development of the cod predators, mentioned in the request, is not taken into account. Reduced predator stocks may enhance the rebuilding of cod, while an increase of predators may inhibit the process and require prolonged strong regulations of the fishery for obtaining the rebuilding target.

### 2.10 Recent ICES advice

Since 2004 the advice has been; No catch should be taken from this stock and a recovery plan should be developed and implemented.

### 2.11 Response to the comments from the Review Group

The SSB values from the VPA runs refer to 1. January
Recent ICES advice is described in section 2.10
A comparison of results with last year's assessment is given in section 2.5.1

Table 2.1a. Norwegian coastal cod. Estimated commercial landings in numbers ('000) at age, and total tonnes by year.

|  | Age |  |  |  |  |  |  |  |  | Tonnes <br> Landed |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |  |
| 1984 | 829 | 3478 | 6954 | 7278 | 6004 | 4964 | 2161 | 819 | 624 | 74824 |
| 1985 | 396 | 7848 | 7367 | 8699 | 7085 | 3066 | 705 | 433 | 264 | 75451 |
| 1986 | 4095 | 4095 | 12662 | 8906 | 5750 | 3868 | 1270 | 342 | 407 | 68905 |
| 1987 | 170 | 940 | 8236 | 12430 | 4427 | 2649 | 1127 | 313 | 149 | 60972 |
| 1988 | 110 | 1921 | 3343 | 6451 | 6626 | 4687 | 1461 | 497 | 333 | 59294 |
| 1989 | 41 | 1159 | 1434 | 2299 | 5197 | 2720 | 949 | 236 | 86 | 40285 |
| 1990 | 7 | 349 | 1233 | 1330 | 1129 | 3456 | 773 | 141 | 73 | 28127 |
| 1991 | 125 | 607 | 1452 | 3114 | 1873 | 1297 | 873 | 132 | 94 | 24822 |
| 1992 | 40 | 665 | 3160 | 4422 | 2992 | 1945 | 898 | 837 | 279 | 41690 |
| 1993 | 4 | 369 | 1706 | 2343 | 2684 | 3072 | 1871 | 627 | 690 | 52557 |
| 1994 | 332 | 573 | 1693 | 4302 | 2467 | 3337 | 1514 | 777 | 798 | 54562 |
| 1995 | 810 | 896 | 2345 | 5188 | 5546 | 3270 | 1455 | 557 | 433 | 57207 |
| 1996 | 1193 | 2376 | 2480 | 4930 | 4647 | 4160 | 2082 | 898 | 543 | 61776 |
| 1997 | 1326 | 3438 | 3150 | 2258 | 2490 | 3935 | 3312 | 959 | 684 | 63319 |
| 1998 | 554 | 2819 | 4786 | 4023 | 2272 | 1546 | 1826 | 975 | 343 | 51572 |
| 1999 | 252 | 1322 | 2346 | 4263 | 2773 | 1602 | 751 | 774 | 320 | 40732 |
| 2000 | 156 | 971 | 3664 | 3807 | 2671 | 1104 | 326 | 132 | 152 | 36715 |
| 2001 | 44 | 505 | 1837 | 2974 | 1998 | 1409 | 542 | 187 | 119 | 29699 |
| 2002 | 192 | 893 | 2331 | 2822 | 2742 | 1538 | 915 | 325 | 377 | 40994 |
| 2003 | 81 | 1107 | 2094 | 2506 | 2158 | 1374 | 598 | 258 | 99 | 34635 |
| 2004 | 12 | 306 | 924 | 1713 | 1820 | 1444 | 609 | 226 | 264 | 24547 |
| 2005 | 15 | 474 | 1299 | 1828 | 1436 | 1115 | 513 | 188 | 143 | 22432 |
| 2006 | 71 | 315 | 1656 | 1695 | 1695 | 1246 | 671 | 326 | 224 | 26134 |
| 2007 | 88 | 515 | 1396 | 1846 | 1252 | 824 | 391 | 256 | 196 | 23841 |
| 2008 | 92 | 670 | 1438 | 1635 | 1232 | 862 | 440 | 215 | 170 | 25777 |
| 2009 | 3 | 238 | 1052 | 1280 | 1388 | 1065 | 545 | 172 | 276 | 24821 |

Table 2.1b. Estimated commercial catch of coastal cod in 2008 by gear and area (tonnes).

| Year |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Area | $\mathbf{2 0 0 9}$ |  |  |  |  |  |
| Gillnet | 03 | 04 | 00 | 05 | $\mathbf{0 6 / 0 7}$ | Total |
| L.linelJig | 832 | 1615 | 3356 | 2475 | 4341 | $\mathbf{1 2 6 1 9} 619$ |
| Danish seine | 1365 | 696 | 1527 | 1390 | 672 | $\mathbf{5 6 5 0}$ |
| Trawl | 912 | 1120 | 987 | 2384 | 304 | $\mathbf{5 7 0 8}$ |
| Total | 385 | 393 | 8 | 46 | 13 | $\mathbf{8 4 4}$ |

Table 2.1c. Norwegian coastal cod. Estimated recreational and tourist catches in numbers ('000) at age, and total tonnes by year.

| Numbers at age |  |  |  |  |  |  |  |  |  |  | Tourist |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | (tonnes) | (tonnes) | (tonnes) |
| 1984 | 650 | 1731 | 2116 | 1667 | 1194 | 597 | 236 | 133 | 13300 | 13300 | 0 |
| 1985 | 3162 | 2590 | 2366 | 1745 | 647 | 225 | 130 | 79 | 13400 | 13400 | 0 |
| 1986 | 627 | 3033 | 2668 | 1659 | 1139 | 435 | 251 | 139 | 13500 | 13500 | 0 |
| 1987 | 108 | 1972 | 4008 | 2181 | 649 | 431 | 109 | 38 | 13500 | 13500 | 0 |
| 1988 | 634 | 1407 | 1567 | 1708 | 2088 | 550 | 129 | 94 | 13600 | 13600 | 0 |
| 1989 | 418 | 825 | 1483 | 1758 | 1413 | 518 | 108 | 34 | 13700 | 13600 | 100 |
| 1990 | 401 | 1494 | 1252 | 682 | 2709 | 450 | 73 | 0 | 14500 | 14400 | 100 |
| 1991 | 1183 | 2698 | 2996 | 1342 | 808 | 583 | 104 | 71 | 15300 | 15200 | 100 |
| 1992 | 429 | 1281 | 2349 | 1491 | 630 | 514 | 846 | 84 | 16100 | 16000 | 100 |
| 1993 | 47 | 1276 | 1288 | 813 | 846 | 696 | 202 | 368 | 14800 | 14700 | 100 |
| 1994 | 57 | 701 | 1723 | 715 | 1288 | 671 | 393 | 124 | 14700 | 14600 | 100 |
| 1995 | 8 | 332 | 804 | 1451 | 1585 | 780 | 413 | 180 | 14700 | 14500 | 200 |
| 1996 | 21 | 591 | 509 | 617 | 1497 | 1373 | 461 | 227 | 14500 | 14300 | 200 |
| 1997 | 51 | 707 | 1023 | 763 | 735 | 1189 | 688 | 132 | 14500 | 14200 | 300 |
| 1998 | 249 | 1137 | 2327 | 1316 | 585 | 410 | 329 | 255 | 14600 | 14300 | 300 |
| 1999 | 49 | 466 | 1445 | 1939 | 920 | 357 | 198 | 221 | 13900 | 13500 | 400 |
| 2000 | 63 | 554 | 1153 | 1515 | 1044 | 344 | 127 | 109 | 13600 | 13100 | 500 |
| 2001 | 0 | 343 | 735 | 1046 | 964 | 873 | 198 | 134 | 13400 | 12700 | 700 |
| 2002 | 56 | 298 | 830 | 1055 | 939 | 596 | 335 | 165 | 13600 | 12800 | 800 |
| 2003 | 85 | 342 | 664 | 916 | 918 | 450 | 244 | 326 | 13900 | 13000 | 900 |
| 2004 | 26 | 254 | 483 | 924 | 1099 | 827 | 358 | 162 | 13400 | 12300 | 1100 |
| 2005 | 21 | 270 | 658 | 858 | 853 | 715 | 423 | 176 | 13200 | 12000 | 1200 |
| 2006 | 19 | 236 | 1016 | 867 | 983 | 612 | 315 | 127 | 13000 | 11700 | 1300 |
| 2007 | 49 | 346 | 759 | 959 | 606 | 531 | 327 | 157 | 13000 | 11500 | 1500 |
| 2008 | 15 | 395 | 743 | 838 | 650 | 400 | 261 | 134 | 12800 | 11200 | 1600 |
| 2009 | 0 | 84 | 576 | 727 | 863 | 600 | 280 | 90 | 12700 | 10900 | 1800 |

Table 2.2. Sampling from cod fisheries in 2009 in the statistical areas $00,03,04,05,06+07$. Number of age samples of cod by quarter, total number of cod otoliths.

| Quarter | 3 | 4 | 0 | 5 | $6+7$ | Tot |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 23 | 36 | 28 | 73 | 24 | 184 |
| 2 | 33 | 18 | 0 | 15 | 6 | 72 |
| 3 | 11 | 2 | 1 | 0 | 0 | 14 |
| 4 | 21 | 28 | 5 | 19 | 16 | 89 |
| Total samples | 88 | 84 | 34 | 107 | 46 | 359 |
| Total otoliths | 2933 | 2765 | 976 | 3404 | 981 | 11059 |
| Coastal cod type otoliths | 492 | 599 | 276 | 508 | 765 | 2640 |

Table 2.3 Number of otoliths sampled by quarter from commercial catches in the period 1985-2009. CC=coastal cod, NEAC=Northeast Arctic cod.

| YEAR | QUARTER | 1 | QUARTER | 2 | QUARTER | 3 | QUARTER | 4 | TOTAL | $\%$ |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | CC | NEAC | CC | NEAC | CC | NEAc | CC | NEAC | CC | NEAC | CC |
| 1985 | 1451 | 3852 | 777 | 1540 | 1277 | 1767 | 1966 | 730 | 5471 | 7889 | 41 |
| 1986 | 940 | 1594 | 1656 | 2579 | 0 | 0 | 669 | 966 | 3265 | 5139 | 39 |
| 1987 | 1195 | 2322 | 937 | 3051 | 638 | 1108 | 1122 | 1137 | 3892 | 7618 | 34 |
| 1988 | 257 | 546 | 160 | 619 | 87 | 135 | 55 | 44 | 559 | 1344 | 29 |
| 1989 | 556 | 1387 | 72 | 374 | 65 | 501 | 97 | 663 | 790 | 2925 | 21 |
| 1990 | 731 | 2974 | 61 | 689 | 252 | 97 | 265 | 674 | 1309 | 4434 | 23 |
| 1991 | 285 | 1168 | 92 | 561 | 77 | 96 | 279 | 718 | 733 | 2543 | 22 |
| 1992 | 152 | 619 | 281 | 788 | 79 | 82 | 272 | 672 | 784 | 2161 | 27 |
| 1993 | 314 | 1098 | 172 | 1046 | 0 | 0 | 310 | 541 | 796 | 2685 | 23 |
| 1994 | 317 | 1605 | 179 | 923 | 21 | 31 | 126 | 674 | 643 | 3233 | 17 |
| 1995 | 188 | 1591 | 232 | 1682 | 2095 | 1057 | 752 | 1330 | 3267 | 5660 | 37 |
| 1996 | 861 | 5486 | 591 | 1958 | 1784 | 1076 | 958 | 2256 | 4194 | 10776 | 28 |
| 1997 | 1106 | 5429 | 367 | 2494 | 1940 | 894 | 1690 | 1755 | 5103 | 10572 | 33 |
| 1998 | 608 | 4930 | 552 | 1342 | 489 | 1094 | 2999 | 2217 | 4648 | 9583 | 33 |
| 1999 | 1277 | 4702 | 493 | 2379 | 202 | 717 | 961 | 1987 | 2933 | 9785 | 23 |
| 2000 | 1283 | 4918 | 365 | 2112 | 386 | 1295 | 472 | 668 | 2506 | 9993 | 20 |
| 2001 | 1102 | 5091 | 352 | 2295 | 126 | 786 | 432 | 983 | 2012 | 9155 | 18 |
| 2002 | 823 | 5818 | 321 | 1656 | 503 | 831 | 897 | 1355 | 2544 | 9660 | 21 |
| 2003 | 821 | 4197 | 445 | 2850 | 790 | 936 | 1112 | 1286 | 3168 | 9269 | 25 |
| 2004 | 1511 | 7539 | 758 | 2565 | 532 | 685 | 531 | 1317 | 3332 | 12106 | 22 |
| 2005 | 1583 | 6219 | 767 | 4383 | 473 | 258 | 877 | 1258 | 3700 | 12188 | 23 |
| 2006 | 2244 | 5087 | 1329 | 2819 | 590 | 271 | 119 | 71 | 4282 | 8248 | 34 |
| 2007 | 1867 | 5895 | 944 | 2496 | 503 | 648 | 637 | 1163 | 3951 | 10202 | 28 |
| 2008 | 1450 | 4162 | 1116 | 3122 | 626 | 515 | 693 | 999 | 3885 | 8798 | 31 |
| 2009 | 1114 | 5109 | 558 | 2592 | 126 | 253 | 842 | 465 | 2640 | 8419 | 24 |

Table 2.4. Landings in tonnes of Coastal cod by area and quarter 2006-2009 (upper 4 tables) Proportion (of total) coastal cod in landings by area and quarter 2006-2009 (lower 4 tables).

| Year | 2006 |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Qu./Area | 03 | 04 | 00 | 05 | $06-07$ | Total |
| $\mathbf{1}$ | 291 | 3483 | 2677 | 3150 | 4169 | 13769 |
| $\mathbf{2}$ | 1485 | 2298 | 601 | 507 | 1388 | 6279 |
| $\mathbf{3}$ | 343 | 893 | 338 | 635 | 564 | 2774 |
| 4 | 253 | 1232 | 444 | 1071 | 312 | 3312 |
|  | Total | 2372 | 7906 | 4059 | 5363 | 6434 |
| 26134 |  |  |  |  |  |  |


| Year |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Qu.IArea | 03 | 04 | 00 | 05 | $06-07$ | Total |
| $\mathbf{1}$ | 664 | 1812 | 3787 | 2274 | 3843 | 12380 |
| $\mathbf{2}$ | 2962 | 1762 | 679 | 803 | 1324 | 7530 |
| $\mathbf{3}$ | 416 | 393 | 537 | 279 | 423 | 2049 |
| $\mathbf{4}$ | 557 | 343 | 346 | 354 | 283 | 1883 |
|  | Total | 4599 | 4311 | 5349 | 3709 | 5873 |


| Year | 2008 |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Qu.IArea | 03 | 04 | 00 | 05 | $06-07$ | Total |
| $\mathbf{1}$ | 653 | 2206 | 3964 | 2222 | 4090 | 13134 |
| $\mathbf{2}$ | 2005 | 2162 | 1116 | 979 | 1640 | 7902 |
| $\mathbf{3}$ | 513 | 647 | 287 | 332 | 434 | 2212 |
| $\mathbf{4}$ | 356 | 793 | 424 | 657 | 299 | 2529 |
| Total | 3526 | 5807 | 5791 | 4190 | 6463 | 25777 |


| Year |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Qu./Area | 03 | 04 | 00 | 05 | $06-07$ | Total |
| $\mathbf{1}$ | 1122 | 1073 | 4537 | 3006 | 3581 | 13318 |
| $\mathbf{2}$ | 723 | 1195 | 715 | 1461 | 985 | 5079 |
| $\mathbf{3}$ | 640 | 394 | 340 | 633 | 398 | 2405 |
| $\mathbf{4}$ | 1009 | 1161 | 286 | 1196 | 367 | 4019 |
|  | Total | 3494 | 3824 | 5877 | 6295 | 5331 |
| 24821 |  |  |  |  |  |  |


| Year |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Qu./Area | $\mathbf{0 3}$ | $\mathbf{0 4}$ | $\mathbf{0 0}$ | $\mathbf{0 5}$ | $\mathbf{0 6 - 0 7}$ | Total |
| $\mathbf{1}$ | 0.05 | 0.20 | 0.13 | 0.13 | 0.88 | $\mathbf{0 . 1 9}$ |
| $\mathbf{2}$ | 0.20 | 0.16 | 0.13 | 0.10 | 0.96 | $\mathbf{0 . 1 9}$ |
| $\mathbf{3}$ | 0.35 | 0.81 | 0.91 | 0.95 | 0.98 | $\mathbf{0 . 7 5}$ |
| $\mathbf{4}$ | 0.10 | 0.85 | 0.91 | 0.95 | 0.99 | $\mathbf{0 . 5 6}$ |
|  | $\mathbf{T o t a l}$ | $\mathbf{0 . 1 5}$ | $\mathbf{0 . 2 3}$ | $\mathbf{0 . 1 5}$ | $\mathbf{0 . 1 7}$ | $\mathbf{0 . 9 1}$ | $\mathbf{0 . 2 3}$|  |
| :--- |


| Year <br> Qu.IArea | $\mathbf{0 3}$ | $\mathbf{0 4}$ | $\mathbf{0 0}$ | $\mathbf{0 5}$ | $\mathbf{0 6 - 0 7}$ | Total |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | 0.08 | 0.09 | 0.24 | 0.07 | 0.79 | $\mathbf{0 . 1 6}$ |
| $\mathbf{2}$ | 0.28 | 0.13 | 0.24 | 0.23 | 0.95 | $\mathbf{0 . 2 3}$ |
| $\mathbf{3}$ | 0.33 | 0.49 | 0.98 | 0.50 | 1.00 | $\mathbf{0 . 5 7}$ |
| $\mathbf{4}$ | 0.23 | 0.36 | 0.98 | 0.52 | 0.90 | $\mathbf{0 . 4 0}$ |
|  | Total | $\mathbf{0 . 2 0}$ | $\mathbf{0 . 1 2}$ | $\mathbf{0 . 2 8}$ | $\mathbf{0 . 1 1}$ | $\mathbf{0 . 8 4}$ |


| Year |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Qu./Area | $\mathbf{0 3}$ | $\mathbf{0 4}$ | $\mathbf{0 0}$ | $\mathbf{0 5}$ | $\mathbf{0 6 - 0 7}$ | Total |
| $\mathbf{1}$ | 0.10 | 0.10 | 0.23 | 0.08 | 0.86 | $\mathbf{0 . 1 7}$ |
| $\mathbf{2}$ | 0.22 | 0.19 | 0.29 | 0.27 | 0.92 | $\mathbf{0 . 2 6}$ |
| $\mathbf{3}$ | 0.30 | 0.60 | 0.95 | 0.60 | 1.00 | $\mathbf{0 . 5 4}$ |
| $\mathbf{4}$ | 0.14 | 0.65 | 0.95 | 0.57 | 1.00 | $\mathbf{0 . 4 4}$ |
|  | $\mathbf{T o t a l}$ | $\mathbf{0 . 1 8}$ | $\mathbf{0 . 1 6}$ | $\mathbf{0 . 2 7}$ | $\mathbf{0 . 1 2}$ | $\mathbf{0 . 8 9}$ | $\mathbf{0 . 2 2}$.


| Year <br> Qu.IArea | $\mathbf{0 3}$ | $\mathbf{0 4}$ | $\mathbf{0 0}$ | $\mathbf{0 5}$ | $\mathbf{0 6 - 0 7}$ | Total |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{1}$ | 0.14 | 0.07 | 0.25 | 0.09 | 0.77 | $\mathbf{0 . 1 7}$ |
| $\mathbf{2}$ | 0.06 | 0.14 | 0.25 | 0.32 | 0.87 | $\mathbf{0 . 1 7}$ |
| $\mathbf{3}$ | 0.25 | 0.35 | 1.00 | 0.81 | 0.98 | $\mathbf{0 . 4 6}$ |
| $\mathbf{4}$ | 0.50 | 0.70 | 0.96 | 0.81 | 0.98 | $\mathbf{0 . 6 9}$ |
|  | Total | $\mathbf{0 . 1 4}$ | $\mathbf{0 . 1 5}$ | $\mathbf{0 . 2 7}$ | $\mathbf{0 . 1 6}$ | $\mathbf{0 . 8 1}$ |

Table 2.5. Coastal cod. Acoustic abundance indices by sub areas and in total in 2009 (in thousands).

| Age (Year class) |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Area | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ |  |
|  | $(08)$ | $(07)$ | $(06)$ | $(05)$ | $(04)$ | $(03)$ | $(02)$ | $(01)$ | $(00)$ | $(99+)$ | Sum |
| 03 | 1356 | 347 | 629 | 736 | 499 | 321 | 174 | 83 | 47 | 39 | 4230 |
| 04 | 1082 | 758 | 954 | 1142 | 630 | 350 | 312 | 167 | 75 | 62 | 5533 |
| 05 | 623 | 146 | 114 | 194 | 95 | 50 | 458 | 39 | 0 | 0 | 1717 |
| 00 | 141 | 52 | 338 | 833 | 716 | 440 | 244 | 302 | 229 | 0 | 3295 |
| 06 | 240 | 742 | 563 | 987 | 556 | 432 | 59 | 202 | 91 | 31 | 3905 |
| 07 | 0 | 13 | 124 | 68 | 39 | 10 | 12 | 0 | 0 | 9 | 275 |
| Total | 3442 | 2059 | 2722 | 3959 | 2536 | 1603 | 1259 | 793 | 443 | 141 | 18955 |

Table 2.6. Coastal cod. Acoustic abundance indices by age 1995 - 2009 (in thousands).

| År | Alder / Age |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ | Sum |
| 1995 | 28707 | 20191 | 13633 | 15636 | 16219 | 9550 | 3174 | 1158 | 781 | 579 | 109628 |
| 1996 | 1756 | 17378 | 22815 | 12382 | 12514 | 6817 | 3180 | 754 | 242 | 5 | 77843 |
| 1997 | 30694 | 18827 | 28913 | 17334 | 12379 | 10612 | 3928 | 1515 | 26 | 663 | 124891 |
| 1998 | 14455 | 13659 | 15003 | 13239 | 7415 | 3137 | 1578 | 315 | 169 | 128 | 69099 |
| 1999 | 6850 | 11309 | 12171 | 10123 | 7197 | 3052 | 850 | 242 | 112 | 54 | 51960 |
| 2000 | 9587 | 11528 | 11612 | 8974 | 7984 | 5451 | 1365 | 488 | 85 | 97 | 57171 |
| 2001 | 8366 | 6729 | 7994 | 7578 | 4751 | 2567 | 1493 | 487 | 189 | 116 | 40270 |
| 2002 | 1329 | 2990 | 4103 | 4940 | 3617 | 2593 | 1470 | 408 | 29 | 128 | 21607 |
| 2003 | 2084 | 2145 | 3545 | 3880 | 2788 | 2389 | 1144 | 589 | 364 | 80 | 19008 |
| 2004 | 3217 | 3541 | 3696 | 4320 | 2758 | 1940 | 783 | 448 | 98 | 110 | 20914 |
| 2005 | 1443 | 1843 | 3525 | 3198 | 3217 | 1700 | 1120 | 552 | 330 | 78 | 17006 |
| 2006 | 1929 | 2525 | 4049 | 3783 | 3472 | 2509 | 1811 | 399 | 229 | 13 | 20719 |
| 2007 | 2202 | 3300 | 4080 | 5518 | 3259 | 2447 | 1444 | 760 | 197 | 34 | 23241 |
| 2008 | 2128 | 2181 | 2475 | 2863 | 2101 | 1219 | 815 | 403 | 319 | 177 | 14681 |
| 2009 | 3442 | 2059 | 2722 | 3959 | 2536 | 1603 | 1259 | 793 | 443 | 141 | 18955 |

Table 2.7. Coastal cod. Mean length (cm) at age 1995-2009.

|  | Age |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ |
| 1995 | 21.5 | 33.0 | 43.0 | 52.0 | 59.1 | 64.1 | 76.0 | 87.4 | 89.0 | 108.3 |
| 1996 | 19.0 | 30.2 | 41.7 | 52.5 | 59.2 | 65.2 | 79.1 | 84.8 | 87.0 | 114.2 |
| 1997 | 16.8 | 28.7 | 40.8 | 51.6 | 58.1 | 65.9 | 73.6 | 80.8 | 102.0 | 110.7 |
| 1998 | 20.3 | 33.3 | 43.8 | 51.4 | 59.1 | 66.3 | 74.1 | 81.0 | 93.2 | 116.9 |
| 1999 | 21.5 | 32.6 | 43.8 | 54.6 | 59.6 | 65.8 | 77.9 | 90.8 | 99.4 | 118.0 |
| 2000 | 21.6 | 33.3 | 43.4 | 53.5 | 61.0 | 66.1 | 75.5 | 90.8 | 99.1 | 105.5 |
| 2001 | 21.1 | 33.3 | 44.5 | 53.6 | 62.9 | 64.7 | 88.7 | 84.2 | 85.7 | 102.1 |
| 2002 | 22.5 | 34.4 | 44.6 | 56.0 | 61.6 | 67.7 | 72.4 | 66.6 | 89.0 | 108.3 |
| 2003 | 18.9 | 33.8 | 42.1 | 51.6 | 60.0 | 67.2 | 72.7 | 76.9 | 84.9 | 94.8 |
| 2004 | 20.7 | 32.9 | 43.5 | 54.5 | 59.9 | 68.0 | 71.9 | 75.0 | 74.6 | 91.8 |
| 2005 | 22.5 | 32.8 | 42.2 | 57.9 | 60.6 | 64.0 | 71.3 | 69.9 | 73.5 | 108.4 |
| 2006 | 22.2 | 36.1 | 47.0 | 55.5 | 61.4 | 68.0 | 69.5 | 77.8 | 87.0 | 100.5 |
| 2007 | 21.6 | 36.0 | 48.0 | 57.9 | 62.2 | 66.8 | 71.8 | 86.6 | 100.2 | 106.3 |
| 2008 | 21.9 | 36.9 | 49.2 | 59.0 | 66.1 | 70.9 | 71.7 | 74.1 | 77.6 | 98.8 |
| 2009 | 20.9 | 34.5 | 47.8 | 57.8 | 65.8 | 70.5 | 77.9 | 78.4 | 85.1 | 73.5 |

Table 2.8. Coastal cod. Mean weight (grams) at age 1995-2009.

|  | Age |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ |
| 1995 | 81 | 390 | 791 | 1525 | 2222 | 2881 | 4665 | 6979 | 6759 | 9897 |
| 1996 | 59 | 252 | 724 | 1433 | 2053 | 2748 | 4722 | 6685 | 6932 | 9723 |
| 1997 | 43 | 240 | 683 | 1364 | 1893 | 2816 | 4426 | 6406 | 7805 | 1827 |
| 1998 | 52 | 372 | 883 | 1456 | 2107 | 2950 | 4319 | 5625 | 8323 | 12468 |
| 1999 | 70 | 323 | 841 | 1675 | 2192 | 2857 | 4540 | 6579 | 9454 | 12902 |
| 2000 | 72 | 365 | 809 | 1554 | 2539 | 3049 | 4352 | 6203 | 8527 | 12066 |
| 2001 | 51 | 396 | 966 | 1524 | 2314 | 3320 | 3695 | 6144 | 8768 | 12468 |
| 2002 | 103 | 428 | 895 | 1741 | 2433 | 3133 | 4273 | 4397 | 7759 | 12992 |
| 2003 | 62 | 385 | 738 | 1353 | 2145 | 3103 | 3981 | 4921 | 6923 | 9956 |
| 2004 | 83 | 352 | 834 | 1690 | 2255 | 3312 | 4150 | 4594 | 4383 | 9733 |
| 2005 | 112 | 359 | 786 | 2168 | 2265 | 2756 | 4174 | 3373 | 4502 | 15887 |
| 2006 | 105 | 474 | 1080 | 1746 | 2430 | 3336 | 3684 | 5125 | 7028 | 14650 |
| 2007 | 103 | 518 | 1185 | 2011 | 2500 | 3160 | 4241 | 6806 | 11051 | 14931 |
| 2008 | 96 | 508 | 1208 | 2095 | 2987 | 3671 | 3976 | 4387 | 5415 | 11588 |
| 2009 | 85 | 434 | 1116 | 2003 | 2894 | 3632 | 4875 | 5400 | 6125 | 4719 |

Table 2.9. Coastal cod. Acoustic biomass indices (tonnes) in 1995-2009.

|  | Age |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ | Sum |
| 1995 | 2337 | 7868 | 10786 | 23846 | 36039 | 27515 | 14445 | 8761 | 4933 | 7779 | 144309 |
| 1996 | 145 | 4386 | 16521 | 17739 | 25687 | 18731 | 15562 | 4376 | 3130 | 46 | 106323 |
| 1997 | 1319 | 4518 | 19748 | 23644 | 23435 | 29884 | 15060 | 8860 | 249 | 8643 | 135360 |
| 1998 | 752 | 5078 | 13247 | 19274 | 15627 | 9255 | 6675 | 1646 | 1329 | 2083 | 74966 |
| 1999 | 477 | 3650 | 10233 | 16960 | 15774 | 8720 | 4723 | 2097 | 1220 | 567 | 64421 |
| 2000 | 688 | 4321 | 9824 | 14464 | 20482 | 17067 | 5936 | 4359 | 926 | 1232 | 79299 |
| 2001 | 425 | 2662 | 7724 | 11548 | 10993 | 8521 | 5517 | 3010 | 1705 | 1917 | 54022 |
| 2002 | 137 | 1279 | 3672 | 8600 | 8801 | 8124 | 6282 | 1794 | 225 | 1663 | 40577 |
| 2003 | 125 | 876 | 2569 | 5328 | 5788 | 6995 | 4201 | 2754 | 2674 | 1136 | 32446 |
| 2004 | 329 | 1269 | 3087 | 7394 | 6089 | 6901 | 3009 | 1779 | 454 | 1058 | 31405 |
| 2005 | 109 | 675 | 2947 | 6521 | 7167 | 4807 | 3648 | 1942 | 1315 | 1205 | 30336 |
| 2006 | 202 | 1197 | 4374 | 6605 | 8435 | 8367 | 6672 | 2045 | 1602 | 190 | 39689 |
| 2007 | 227 | 1709 | 4835 | 11097 | 8148 | 7733 | 6124 | 5173 | 2177 | 508 | 47731 |
| 2008 | 206 | 1212 | 3120 | 6085 | 6593 | 4203 | 3437 | 2014 | 1492 | 2066 | 30506 |
| 2009 | 294 | 893 | 3037 | 7933 | 7335 | 5821 | 6137 | 4282 | 2707 | 665 | 39107 |

Table 2.10. Coastal cod. Maturity ogives by age in the period 1995-2009.

|  |  |  |  |  | Age |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ |
| 1995 | 0.00 | 0.00 | 0.01 | 0.21 | 0.48 | 0.71 | 0.87 | 0.87 | 1.00 | 1.00 |
| 1996 | 0.00 | 0.00 | 0.03 | 0.25 | 0.56 | 0.81 | 0.92 | 0.99 | 1.00 | 1.00 |
| 1997 | 0.00 | 0.00 | 0.06 | 0.29 | 0.45 | 0.76 | 0.97 | 1.00 | 1.00 | 1.00 |
| 1998 | 0.00 | 0.02 | 0.15 | 0.25 | 0.53 | 0.74 | 0.87 | 0.89 | 1.00 | 1.00 |
| 1999 | 0.00 | 0.02 | 0.03 | 0.21 | 0.43 | 0.66 | 0.74 | 1.00 | 1.00 | 1.00 |
| 2000 | 0.00 | 0.00 | 0.00 | 0.16 | 0.31 | 0.61 | 0.76 | 0.64 | 0.99 | 1.00 |
| 2001 | 0.00 | 0.00 | 0.00 | 0.04 | 0.37 | 0.78 | 0.98 | 0.99 | 0.97 | 1.00 |
| 2002 | 0.00 | 0.02 | 0.02 | 0.26 | 0.88 | 0.93 | 0.90 | 0.97 | 1.00 | 1.00 |
| 2003 | 0.00 | 0.00 | 0.00 | 0.05 | 0.29 | 0.49 | 0.90 | 0.98 | 0.96 | 1.00 |
| 2004 | 0.00 | 0.00 | 0.01 | 0.09 | 0.37 | 0.76 | 0.95 | 0.98 | 1.00 | 1.00 |
| 2005 | 0.00 | 0.00 | 0.00 | 0.07 | 0.40 | 0.56 | 0.89 | 0.98 | 1.00 | 1.00 |
| 2006 | 0.00 | 0.00 | 0.00 | 0.14 | 0.52 | 0.75 | 0.91 | 0.87 | 0.96 | 1.00 |
| 2007 | 0.00 | 0.00 | 0.00 | 0.14 | 0.54 | 0.76 | 0.96 | 0.83 | 1.00 | 1.00 |
| 2008 | 0.00 | 0.00 | 0.03 | 0.12 | 0.48 | 0.72 | 0.89 | 0.94 | 0.96 | 1.00 |
| 2009 | 0.00 | 0.00 | 0.02 | 0.06 | 0.26 | 0.35 | 0.59 | 0.74 | 0.60 | 0.92 |

Table 2.11. Coastal cod. Acoustic spawning biomass indices (tonnes) in 1995 - 2009.

|  | Age |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ | Sum |
| 1995 | 0 | 0 | 96 | 4925 | 17424 | 19614 | 12573 | 7648 | 4933 | 7779 | 74992 |
| 1996 | 0 | 0 | 468 | 4467 | 14320 | 15130 | 14365 | 4311 | 3130 | 46 | 56237 |
| 1997 | 0 | 0 | 1185 | 6857 | 10546 | 22712 | 14608 | 8860 | 249 | 8643 | 73660 |
| 1998 | 0 | 92 | 2026 | 4870 | 8252 | 6804 | 5774 | 1461 | 1329 | 2083 | 32691 |
| 1999 | 0 | 56 | 315 | 3544 | 6778 | 5716 | 3478 | 2097 | 1220 | 567 | 23771 |
| 2000 | 0 | 0 | 0 | 2366 | 6354 | 10426 | 4486 | 2798 | 916 | 1232 | 28579 |
| 2001 | 0 | 0 | 15 | 508 | 4102 | 6662 | 5398 | 2978 | 1650 | 1917 | 23230 |
| 2002 | 0 | 20 | 87 | 2240 | 7702 | 7551 | 5650 | 1747 | 225 | 1663 | 26885 |
| 2003 | 0 | 0 | 0 | 269 | 1670 | 3428 | 3778 | 2686 | 2554 | 1136 | 15521 |
| 2004 | 0 | 0 | 28 | 679 | 2252 | 5253 | 2853 | 1736 | 434 | 722 | 13959 |
| 2005 | 0 | 0 | 0 | 447 | 2844 | 2670 | 3247 | 1898 | 1315 | 288 | 12709 |
| 2006 | 0 | 0 | 0 | 925 | 4386 | 6275 | 6072 | 1779 | 1538 | 571 | 21546 |
| 2007 | 0 | 0 | 0 | 1554 | 4400 | 5877 | 5879 | 4294 | 2177 | 508 | 24689 |
| 2008 | 0 | 0 | 107 | 734 | 3189 | 3012 | 3049 | 1902 | 1434 | 2066 | 15493 |
| 2009 | 0 | 0 | 61 | 476 | 1907 | 2037 | 3621 | 3169 | 1624 | 612 | 13508 |

Table 2.12. Proportion coastal cod among sampled cod during the coastal survey by age and statistical areas in the years 2004-2009.

| Year | Area/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 3 | 0,61 | 0,62 | 0,35 | 0,43 | 0,39 | 0,34 | 0,45 | 0,33 | 0,69 |
| 2004 | 4 | 0,84 | 0,83 | 0,74 | 0,76 | 0,77 | 0,47 | 0,77 | 0,44 | 0,44 |
| 2004 | 5 | 0,80 | 0,89 | 0,82 | 0,79 | 0,62 | 0,85 | 0,75 | 0,50 | 0,20 |
| 2004 | 0 | 1,00 | 0,94 | 0,94 | 0,60 | 0,85 | 1,00 | 1,00 | 1,00 | 0,07 |
| 2004 | 6 | 0,85 | 0,94 | 0,86 | 0,85 | 0,74 | 0,77 | 0,64 | 1,00 |  |
| 2004 | 7 | 0,98 | 0,96 | 0,99 | 0,97 | 0,90 | 0,91 | 0,75 | 1,00 |  |
| 2005 | 3 | 0,63 | 0,54 | 0,54 | 0,45 | 0,35 | 0,30 | 0,20 | 0,48 | 0,03 |
| 2005 | 4 | 0,96 | 0,91 | 0,76 | 0,74 | 0,71 | 0,60 | 0,76 | 0,81 | 0,50 |
| 2005 | 5 | 0,00 | 0,54 | 0,65 | 0,68 | 0,52 | 1,00 | 1,00 | 0,67 |  |
| 2005 | 0 | 0,11 | 0,39 | 0,70 | 0,61 | 0,70 | 0,85 | 0,50 | 1,00 |  |
| 2005 | 6 | 1,00 | 1,00 | 0,93 | 0,87 | 0,81 | 0,81 | 0,59 | 0,96 |  |
| 2005 | 7 | 1,00 | 1,00 | 1,00 | 1,00 | 1,00 | 0,86 | 0,67 | 0,00 |  |
| 2006 | 3 | 0,79 | 0,77 | 0,63 | 0,59 | 0,45 | 0,37 | 0,30 | 0,39 | 0,00 |
| 2006 | 4 | 1,00 | 0,88 | 0,84 | 0,79 | 0,68 | 0,63 | 0,82 | 0,40 | 0,42 |
| 2006 | 5 | 1,00 | 0,98 | 0,81 | 0,88 | 0,77 | 0,63 | 0,80 | 0,00 | 0,50 |
| 2006 | 0 | 0,99 | 0,99 | 0,95 | 0,87 | 0,86 | 0,89 | 0,85 | 0,33 |  |
| 2006 | 6 | 1,00 | 1,00 | 0,95 | 0,99 | 0,80 | 0,72 | 1,00 | 0,67 |  |
| 2006 | 7 | 1,00 | 0,97 | 0,95 | 0,98 | 0,89 | 1,00 | 0,50 |  |  |
| 2007 | 3 | 0,83 | 0,38 | 0,40 | 0,59 | 0,27 | 0,32 | 0,00 | 1,00 |  |
| 2007 | 4 | 0,91 | 0,92 | 0,92 | 0,80 | 0,80 | 0,90 | 0,71 | 0,67 | 1,00 |
| 2007 | 5 | 0,97 | 1,00 | 0,97 | 0,94 | 0,94 | 0,95 | 0,86 | 0,67 | 0,00 |
| 2007 | 0 | 1,00 | 0,88 | 1,00 | 1,00 | 1,00 | 0,00 | 1,00 | 1,00 |  |
| 2007 | 6 | 1,00 | 1,00 | 0,95 | 0,87 | 0,91 | 0,81 |  |  |  |
| 2007 | 7 | 1,00 | 1,00 | 1,00 | 0,89 | 0,86 | 0,86 | 1,00 | 1,00 | 1,00 |
| 2008 | 3 | 0.98 | 0.97 | 0.80 | 0.83 | 0.79 | 0.72 | 0.53 | 1.00 | 0.40 |
| 2008 | 4 | 1.00 | 0.99 | 0.80 | 0.88 | 0.84 | 0.78 | 0.88 | 0.88 | 0.86 |
| 2008 | 5 | 1.00 | 1.00 | 0.93 | 0.96 | 1.00 | 0.80 | 0.67 | 1.00 | 1.00 |
| 2008 | 0 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.00 |
| 2008 | 6 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2008 | 7 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2009 | 3 | 0.90 | 0.72 | 0.54 | 0.44 | 0.48 | 0.57 | 0.79 | 0.67 | 0.58 |
| 2009 | 4 | 0.95 | 0.89 | 0.78 | 0.62 | 0.69 | 0.92 | 0.72 | 0.78 | 0.79 |
| 2009 | 5 | 1.00 | 1.00 | 0.95 | 0.84 | 0.78 | 0.82 | 0.88 | 0.67 | 1.00 |
| 2009 | 0 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.50 | 1.00 |  |
| 2009 | 6 | 1.00 | 1.00 | 1.00 | 1.00 | 0.82 | 1.00 | 1.00 | 1.00 | 0.50 |
| 2009 | 7 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |  | 1.00 |

Table 2.13. Norwegian Coastal Cod. Input data to all the VPA-analysis.

Table 2 Catch weights at age (kg)

| YEAR | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 2 | 0.248 | 0.214 | 0.227 | 0.331 | 0.246 | 0.3 |  |  |  |  |
| 3 | 0.619 | 0.712 | 0.525 | 0.673 | 0.634 | 0.661 |  |  |  |  |
| 4 | 1.149 | 1.415 | 1.08 | 1.12 | 1.17 | 1.836 |  |  |  |  |
| 5 | 1.734 | 2.036 | 1.706 | 1.693 | 1.727 | 2.17 |  |  |  |  |
| 6 | 2.325 | 2.737 | 2.256 | 2.359 | 2.328 | 2.448 |  |  |  |  |
| 7 | 3.486 | 4.012 | 3.353 | 3.743 | 3.256 | 4.391 |  |  |  |  |
| 8 | 4.845 | 6.116 | 4.838 | 5.326 | 4.7 | 4.899 |  |  |  |  |
| 9 | 5.608 | 6.46 | 5.838 | 6.129 | 5.45 | 6.661 |  |  |  |  |
| +gp | 8.84 | 10.755 | 7.053 | 11.623 | 8.202 | 11.608 |  |  |  |  |
| 0 SOPCOFAC | 1.0002 | 1 | 1.0001 | 1.0001 | 1.0001 | 1 |  |  |  |  |
| YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |

AGE

| 2 | 0.345 | 0.164 | 0.168 | 0.241 | 0.254 | 0.302 | 0.274 | 0.277 | 0.376 | 0.467 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 1.174 | 0.922 | 0.556 | 0.645 | 0.805 | 0.71 | 0.921 | 0.97 | 0.978 | 1.155 |
| 4 | 1.515 | 1.608 | 1.359 | 1.71 | 1.476 | 1.335 | 1.464 | 1.554 | 1.518 | 1.633 |
| 5 | 1.678 | 2.108 | 2.267 | 2.591 | 2.097 | 1.842 | 1.979 | 1.97 | 2.281 | 2.171 |
| 6 | 2.708 | 2.507 | 2.957 | 3.588 | 3.287 | 2.467 | 2.516 | 2.897 | 3.125 | 3.249 |
| 7 | 3.898 | 3.469 | 3.903 | 4.366 | 4.095 | 4.191 | 3.461 | 3.716 | 3.9 | 4.095 |
| 8 |  | 6.515 | 4.976 | 5.317 | 5.899 | 5.592 | 5.778 | 4.866 | 4.829 | 5.52 |
| 9 |  | 7.299 | 5.734 | 4.558 | 6.494 | 7.217 | 6.376 | 5.391 | 6.349 | 6.333 |
|  |  | 13.924 | 11.059 | 7.032 | 7.509 | 8.331 | 9.903 | 8.854 | 9.267 | 9.337 |
|  | +gp | 1.0002 | 1.0003 | 1.0001 | 1 | 1 | 1.0001 | 1.0001 | 1.0003 | 0.9919 |
| 0 | SOPCOFAC | 1.0002 |  |  |  |  |  |  |  |  |
|  | YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
|  |  |  |  |  |  |  |  |  |  |  |

AGE

| 2 | 0.515 | 0.164 | 0.491 | 0.944 | 0.824 | 0.82 | 1.274 | 1.241 | 0.977 | 1.219 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 1.305 | 0.952 | 1.179 | 1.552 | 1.374 | 1.317 | 1.599 | 1.744 | 1.882 | 1.47 |
| 4 | 2.272 | 1.637 | 1.8 | 2.146 | 1.877 | 2.094 | 1.894 | 2.143 | 2.444 | 2.348 |
| 5 | 2.555 | 2.881 | 2.485 | 3.082 | 2.679 | 2.795 | 2.687 | 2.718 | 3.747 | 3.331 |
| 6 | 3.283 | 3.424 | 3.86 | 3.594 | 3.365 | 3.493 | 3.562 | 4.098 | 4.165 | 4.251 |
| 7 | 4.504 | 4.038 | 4.76 | 4.953 | 4.013 | 4.087 | 4.029 | 4.884 | 4.989 | 4.824 |
| 8 | 5.4 | 5.397 | 5.195 | 5.736 | 4.847 | 4.836 | 5.182 | 5.939 | 5.992 | 5.807 |
| 9 | 6.379 | 7.208 | 5.507 | 6.477 | 5.554 | 6.264 | 5.905 | 6.89 | 6.143 | 6.776 |
|  |  | 6.42 | 6.881 | 9.183 | 9.686 | 6.343 | 5.115 | 6.213 | 8.098 | 8.229 |
| 8.5971 |  |  |  |  |  |  |  |  |  |  |
| 0 | +gp | 0.9999 | 1.0004 | 1.0181 | 1.0001 | 0.9997 | 1.0001 | 0.9999 | 0.9998 | 0.9999 |

Table 2.13. Norwegian Coastal Cod. Input data to all the VPA-analysis. Continued

Table 3 Stock weights at age (kg)
$\begin{array}{llllllll}\text { YEAR } & 1984 & 1985 & 1986 & 1987 & 1988 & 1989\end{array}$

AGE

| 2 | 0.321 | 0.321 | 0.321 | 0.321 | 0.321 | 0.321 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.758 | 0.758 | 0.758 | 0.758 | 0.758 | 0.758 |
| 4 | 1.479 | 1.479 | 1.479 | 1.479 | 1.479 | 1.479 |
| 5 | 2.137 | 2.137 | 2.137 | 2.137 | 2.137 | 2.137 |
| 6 | 2.814 | 2.814 | 2.814 | 2.814 | 2.814 | 2.814 |
| 7 | 4.722 | 4.722 | 4.722 | 4.722 | 4.722 | 4.722 |
| 8 | 6.685 | 6.685 | 6.685 | 6.685 | 6.685 | 6.685 |
| 9 | 6.98 | 6.98 | 6.98 | 6.98 | 6.98 | 6.98 |
| +gp | 9.723 | 9.723 | 9.723 | 9.723 | 9.723 | 9.723 |

$\begin{array}{lllllllllllll}\text { YEAR } & 1990 & 1991 & 1992 & 1993 & 1994 & 1995 & 1996 & 1997 & 1998 & 1999\end{array}$

AGE

| 2 | 0.321 | 0.321 | 0.321 | 0.321 | 0.321 | 0.298 | 0.27 | 0.232 | 0.323 | 0.318 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 0.758 | 0.758 | 0.758 | 0.758 | 0.758 | 0.7 | 0.717 | 0.677 | 0.834 | 0.804 |
| 4 | 1.479 | 1.479 | 1.479 | 1.479 | 1.479 | 1.338 | 1.435 | 1.363 | 1.366 | 1.559 |
| 5 | 2.137 | 2.137 | 2.137 | 2.137 | 2.137 | 1.973 | 2.044 | 1.903 | 2.075 | 2.042 |
| 6 | 2.814 | 2.814 | 2.814 | 2.814 | 2.814 | 2.649 | 2.694 | 2.816 | 3.013 | 2.798 |
| 7 | 4.722 | 4.722 | 4.722 | 4.722 | 4.722 | 4.164 | 4.817 | 3.833 | 4.255 | 4.678 |
| 8 | 6.685 | 6.685 | 6.685 | 6.685 | 6.685 | 7.051 | 6.28 | 5.849 | 5.305 | 7.151 |
| 9 | 6.98 | 6.98 | 6.98 | 6.98 | 6.98 | 6.413 | 11.365 | 9.6 | 8.35 | 8.959 |
|  |  | 9.723 | 9.723 | 9.723 | 9.723 | 9.723 | 14.326 | 15.67 | 13.037 | 18.016 | 18.34

$\begin{array}{lllllllllllll}\text { YEAR } & 2000 & 2001 & 2002 & 2003 & 2004 & 2005 & 2006 & 2007 & 2008 & 2009\end{array}$

AGE

| 2 | 0.346 | 0.347 | 0.43 | 0.308 | 0.339 | 0.407 | 0.49 | 0.518 | 0.508 | 0.434 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 0.777 | 0.878 | 0.88 | 0.686 | 0.834 | 0.846 | 1.125 | 1.185 | 1.208 | 1.116 |
| 4 | 1.458 | 1.543 | 1.698 | 1.299 | 1.614 | 1.748 | 1.812 | 2.011 | 2.095 | 2.003 |
| 5 | 2.296 | 2.213 | 2.452 | 2.149 | 2.269 | 2.2 | 2.559 | 2.5 | 2.987 | 2.894 |
| 6 | 2.735 | 2.862 | 3.538 | 3.135 | 3.29 | 2.693 | 3.579 | 3.16 | 3.671 | 3.632 |
| 7 | 4.048 | 3.321 | 4.397 | 4.048 | 4.124 | 3.817 | 3.964 | 4.241 | 3.976 | 4.875 |
| 8 | 7.011 | 4.849 | 4.191 | 5.008 | 4.718 | 3.797 | 4.822 | 6.806 | 4.387 | 5.4 |
| 9 |  | 9.224 | 7.339 | 7.046 | 5.789 | 4.976 | 5.344 | 7.332 | 11.051 | 5.415 |
|  | + gp | 12.277 | 11.542 | 15.619 | 10.069 | 6.358 | 14.829 | 14.65 | 14.931 | 11.558 |

Table 2.13. Norwegian Coastal Cod. Input data to all the VPA-analysis. Continued

Table 5 Proportion mature at age

| YEAR | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |

AGE

| 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 |
| 4 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 |
| 5 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 |
| 6 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 |
| 7 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 |
| 8 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 |
| 9 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 |
|  |  | 1 | 1 | 1 | 1 | 1 |


| YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

AGE

| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 |
| 4 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 |
| 5 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 |
| 6 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 |
| 7 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 |
| 8 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 |
| 9 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 |
|  |  | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |


| YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

AGE

| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 |
| 4 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 |
| 5 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 | 0.46 |
| 6 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 |
| 7 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 | 0.87 |
| 8 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 | 0.91 |
| 9 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 |
|  |  | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

Table 2.14. Norwegian Coastal Cod. Catch numbers at age (thousands) and total catch $i$ (tones) as input to the VPA-analysis including recreational and tourist fisheries.

| Table 1 Catch numbers at age |  |  |  |  |  | Numbers*10**-3 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
| 2 |  | 1479 | 3558 | 4722 | 278 | 744 | 459 |  |  |  |  |
| 3 |  | 5209 | 10438 | 7128 | 2912 | 3328 | 1984 |  |  |  |  |
| 4 |  | 9070 | 9733 | 15330 | 12244 | 4910 | 2917 |  |  |  |  |
| 5 |  | 8945 | 10444 | 10565 | 14611 | 8159 | 4057 |  |  |  |  |
| 6 |  | 7198 | 7732 | 6889 | 5076 | 8714 | 6610 |  |  |  |  |
| 7 |  | 5561 | 3291 | 4303 | 3080 | 5237 | 3238 |  |  |  |  |
| 8 |  | 2397 | 835 | 1521 | 1236 | 1590 | 1057 |  |  |  |  |
| 9 |  | 952 | 512 | 481 | 351 | 591 | 270 |  |  |  |  |
|  | +gp | 624 | 264 | 407 | 149 | 333 | 86 |  |  |  |  |
| 0 | TOTALNUM | 41435 | 46807 | 51346 | 39937 | 33606 | 20678 |  |  |  |  |
|  | TONSLAND | 88124 | 88851 | 82405 | 74472 | 72894 | 53985 |  |  |  |  |
|  | SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 |  |  |  |  |
|  | YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
| 2 |  | 408 | 1308 | 469 | 51 | 389 | 818 | 1214 | 1377 | 803 | 301 |
| 3 |  | 1843 | 3305 | 1946 | 1645 | 1274 | 1228 | 2967 | 4145 | 3956 | 1788 |
| 4 |  | 2485 | 4448 | 5509 | 2994 | 3416 | 3149 | 2989 | 4173 | 7113 | 3791 |
| 5 |  | 2012 | 4456 | 5913 | 3156 | 5017 | 6639 | 5547 | 3021 | 5339 | 6202 |
| 6 |  | 3838 | 2681 | 3622 | 3530 | 3755 | 7131 | 6144 | 3225 | 2857 | 3693 |
| 7 |  | 3906 | 1880 | 2459 | 3768 | 4008 | 4050 | 5533 | 5124 | 1956 | 1959 |
| 8 |  | 846 | 977 | 1744 | 2073 | 1907 | 1868 | 2543 | 4000 | 2155 | 949 |
| 9 |  | 141 | 203 | 921 | 995 | 901 | 737 | 1125 | 1091 | 1230 | 995 |
|  | +gp | 73 | 94 | 279 | 690 | 798 | 433 | 543 | 684 | 343 | 320 |
| 0 | TOTALNUM | 15552 | 19352 | 22862 | 18902 | 21465 | 26053 | 28605 | 26840 | 25752 | 19998 |
|  | TONSLAND | 42627 | 40122 | 57790 | 67357 | 69262 | 71907 | 76276 | 77819 | 66172 | 54632 |
|  | SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 99 | 100 |
|  | YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
| 2 |  | 219 | 44 | 248 | 166 | 38 | 36 | 90 | 137 | 107 | 3 |
| 3 |  | 1525 | 848 | 1191 | 1449 | 560 | 744 | 551 | 861 | 1065 | 322 |
| 4 |  | 4817 | 2572 | 3161 | 2758 | 1407 | 1957 | 2672 | 2155 | 2181 | 1628 |
| 5 |  | 5322 | 4020 | 3877 | 3422 | 2637 | 2686 | 2562 | 2805 | 2473 | 2007 |
| 6 |  | 3715 | 2962 | 3681 | 3076 | 2919 | 2289 | 2678 | 1858 | 1882 | 2251 |
| 7 |  | 1448 | 2282 | 2134 | 1824 | 2271 | 1830 | 1858 | 1355 | 1262 | 1665 |
| 8 |  | 453 | 740 | 1250 | 842 | 967 | 936 | 986 | 718 | 701 | 825 |
| 9 |  | 241 | 321 | 490 | 584 | 388 | 364 | 453 | 413 | 349 | 262 |
|  | +gp | 152 | 119 | 377 | 99 | 264 | 143 | 224 | 196 | 170 | 276 |
| 0 | TOTALNUM | 17892 | 13908 | 16409 | 14220 | 11451 | 10985 | 12074 | 10498 | 10190 | 9239 |
|  | TONSLAND | 50315 | 43099 | 54594 | 48535 | 37947 | 35632 | 39134 | 36841 | 38577 | 37521 |
|  | SOPCOF \% | 100 | 100 | 101 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |

Table 2.15. Norwegian Coastal Cod. Diagnostic output from XSA run for 2009 updated.

| Lowestoft VPA Version 3.1 |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 26/04/2010 15:29 |  |  |  |  |  |  |  |  |  |  |
| Extended Survivors Analysis |  |  |  |  |  |  |  |  |  |  |
| Norwegian Coastal C | COMBSEX | PLUSGROUP |  |  |  |  |  |  |  |  |
| CPUE data from file coast-9.txt |  |  |  |  |  |  |  |  |  |  |
| Catch data for 26 years. 1984 to 2009. Ages 2 to 10. |  |  |  |  |  |  |  |  |  |  |
| Fleet |  |  |  | Last | Alpha | Beta |  |  |  |  |
|  | year | year | age | age |  |  |  |  |  |  |
| Norw. Coast. survey | 1995 | 2009 | 0 | 8 | 0.75 | 0.85 |  |  |  |  |
| Time series weights : |  |  |  |  |  |  |  |  |  |  |
| Tapered time weighting applied |  |  |  |  |  |  |  |  |  |  |
| Power = 3 over 20 years |  |  |  |  |  |  |  |  |  |  |
| Catchability analysis : |  |  |  |  |  |  |  |  |  |  |
| Catchability independent of stock size for all ages |  |  |  |  |  |  |  |  |  |  |
| Catchability independent of age for ages $>=8$ |  |  |  |  |  |  |  |  |  |  |
| Terminal population estimation : |  |  |  |  |  |  |  |  |  |  |
| Survivor estimates shrunk towards the mean F |  |  |  |  |  |  |  |  |  |  |
| of the final 2 years or the 4 oldest ages. |  |  |  |  |  |  |  |  |  |  |
| S.E. of the mean to which the estimates are shrunk $=1.000$ |  |  |  |  |  |  |  |  |  |  |
| Minimum standard error for population |  |  |  |  |  |  |  |  |  |  |
| estimates derived from each fleet $=.300$ |  |  |  |  |  |  |  |  |  |  |
| Prior weighting not applied |  |  |  |  |  |  |  |  |  |  |
| Tuning had not converged after 30 iterations |  |  |  |  |  |  |  |  |  |  |
| Total absolute residual between iterations |  |  |  |  |  |  |  |  |  |  |
| 29 and 30 = . 00525 |  |  |  |  |  |  |  |  |  |  |
| Final year F values |  |  |  |  |  |  |  |  |  |  |
| Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |  |  |
| Iteration 29 | 0.001 | 0.0426 | 0.1797 | 0.3381 | 0.5365 | 0.6032 | 0.4827 | 0.3102 |  |  |
| Iteration 30 | 0.001 | 0.0426 | 0.1795 | 0.3378 | 0.5359 | 0.6021 | 0.4808 | 0.3089 |  |  |
| Regression weights |  |  |  |  |  |  |  |  |  |  |
|  | 0.751 | 0.82 | 0.877 | 0.921 | 0.954 | 0.976 | 0.99 | 0.997 | 1 | 1 |
| Fishing mortalities |  |  |  |  |  |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 2 | 0.008 | 0.002 | 0.012 | 0.005 | 0.001 | 0.001 | 0.006 | 0.008 | 0.013 | 0.001 |
| 3 | 0.054 | 0.031 | 0.059 | 0.087 | 0.025 | 0.038 | 0.03 | 0.059 | 0.082 | 0.043 |
| 4 | 0.243 | 0.138 | 0.193 | 0.191 | 0.098 | 0.139 | 0.18 | 0.182 | 0.234 | 0.18 |
| 5 | 0.396 | 0.318 | 0.327 | 0.328 | 0.236 | 0.285 | 0.272 | 0.313 | 0.337 | 0.338 |
| 6 | 0.471 | 0.373 | 0.548 | 0.447 | 0.422 | 0.318 | 0.468 | 0.332 | 0.357 | 0.536 |
| 7 | 0.442 | 0.49 | 0.554 | 0.591 | 0.619 | 0.498 | 0.505 | 0.437 | 0.402 | 0.602 |
| 8 | 0.304 | 0.406 | 0.695 | 0.434 | 0.573 | 0.464 | 0.644 | 0.29 | 0.443 | 0.481 |
| 9 | 0.314 | 0.286 | 0.457 | 0.424 | 0.288 | 0.345 | 0.614 | 0.547 | 0.256 | 0.309 |

XSA population numbers (Thousands)

|  | AGE |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| YEAR | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| 2000 | $2.28 \mathrm{E}+04$ | $2.03 \mathrm{E}+04$ | $1.88 \mathrm{E}+04$ | $1.29 \mathrm{E}+04$ | $7.87 \mathrm{E}+03$ | $3.41 \mathrm{E}+03$ | $1.38 \mathrm{E}+03$ | $5.42 \mathrm{E}+02$ |
| 2001 | $2.11 \mathrm{E}+04$ | $1.85 \mathrm{E}+04$ | $1.57 \mathrm{E}+04$ | $1.21 \mathrm{E}+04$ | $7.09 \mathrm{E}+03$ | $4.02 \mathrm{E}+03$ | $1.80 \mathrm{E}+03$ | $8.31 \mathrm{E}+02$ |
| 2002 | $1.81 \mathrm{E}+04$ | $1.73 \mathrm{E}+04$ | $1.47 \mathrm{E}+04$ | $1.12 \mathrm{E}+04$ | $7.19 \mathrm{E}+03$ | $3.99 \mathrm{E}+03$ | $2.02 \mathrm{E}+03$ | $9.80 \mathrm{E}+02$ |
| 2003 | $1.70 \mathrm{E}+04$ | $1.46 \mathrm{E}+04$ | $1.33 \mathrm{E}+04$ | $9.91 \mathrm{E}+03$ | $6.61 \mathrm{E}+03$ | $3.40 \mathrm{E}+03$ | $1.88 \mathrm{E}+03$ | $8.25 \mathrm{E}+02$ |
| 2004 | $1.72 \mathrm{E}+04$ | $1.38 \mathrm{E}+04$ | $1.10 \mathrm{E}+04$ | $9.01 \mathrm{E}+03$ | $5.85 \mathrm{E}+03$ | $3.46 \mathrm{E}+03$ | $1.54 \mathrm{E}+03$ | $9.97 \mathrm{E}+02$ |
| 2005 | $1.43 \mathrm{E}+04$ | $1.41 \mathrm{E}+04$ | $1.10 \mathrm{E}+04$ | $8.14 \mathrm{E}+03$ | $5.83 \mathrm{E}+03$ | $3.14 \mathrm{E}+03$ | $1.53 \mathrm{E}+03$ | $7.12 \mathrm{E}+02$ |
| 2006 | $1.21 \mathrm{E}+04$ | $1.17 \mathrm{E}+04$ | $1.11 \mathrm{E}+04$ | $7.86 \mathrm{E}+03$ | $5.01 \mathrm{E}+03$ | $3.47 \mathrm{E}+03$ | $1.56 \mathrm{E}+03$ | $7.85 \mathrm{E}+02$ |
| 2007 | $1.16 \mathrm{E}+04$ | $9.87 \mathrm{E}+03$ | $9.27 \mathrm{E}+03$ | $7.58 \mathrm{E}+03$ | $4.90 \mathrm{E}+03$ | $2.57 \mathrm{E}+03$ | $1.71 \mathrm{E}+03$ | $6.71 \mathrm{E}+02$ |
| 2008 | $7.81 \mathrm{E}+03$ | $9.38 \mathrm{E}+03$ | $7.62 \mathrm{E}+03$ | $6.32 \mathrm{E}+03$ | $4.54 \mathrm{E}+03$ | $2.88 \mathrm{E}+03$ | $1.36 \mathrm{E}+03$ | $1.05 \mathrm{E}+03$ |
| 2009 | $3.21 \mathrm{E}+03$ | $6.31 \mathrm{E}+03$ | $7.07 \mathrm{E}+03$ | $4.93 \mathrm{E}+03$ | $3.70 \mathrm{E}+03$ | $2.60 \mathrm{E}+03$ | $1.58 \mathrm{E}+03$ | $7.15 \mathrm{E}+02$ |
| Estimated population abundance at 1 st Jan 2010 |  |  |  |  |  |  |  |  |

$0.00 \mathrm{E}+00 \quad 2.62 \mathrm{E}+03 \quad 4.95 \mathrm{E}+03 \quad 4.84 \mathrm{E}+03 \quad 2.88 \mathrm{E}+03 \quad 1.77 \mathrm{E}+03 \quad 1.17 \mathrm{E}+03 \quad 8.02 \mathrm{E}+02$
Taper weighted geometric mean of the VPA populations:
$\begin{array}{lllllll}1.54 \mathrm{E}+04 & 1.49 \mathrm{E}+04 & 1.29 \mathrm{E}+04 & 9.71 \mathrm{E}+03 & 6.37 \mathrm{E}+03 & 3.75 \mathrm{E}+03 & 1.94 \mathrm{E}+03 \\ 9.85 \mathrm{E}+02\end{array}$
Standard error of the weighted $\log ($ VPA populations) :

| 0.6542 | 0.4597 | 0.3982 | 0.3963 | 0.3799 | 0.407 | 0.4436 | 0.4498 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

1
Fleet: Norw. Coast. survey

| Age | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.76 | 0.44 | 0.74 | 0.46 | 0.46 |  |  |  |  |  |
| 3 | 0.57 | 0.85 | 0.94 | 0.5 | 0.28 |  |  |  |  |  |
| 4 | 0.54 | 0.57 | 0.69 | 0.33 | 0.19 |  |  |  |  |  |
| 5 | 0.33 | 0.84 | 0.92 | 0.31 | 0.2 |  |  |  |  |  |
| 6 | -0.01 | -0.02 | 1.32 | 0.13 | 0.09 |  |  |  |  |  |
| 7 | -0.13 | -0.42 | 0.29 | 0.29 | -0.26 |  |  |  |  |  |
| 8 | -0.13 | -0.44 | 0.1 | -0.97 | -0.31 |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 2 | 0.57 | 0.11 | -0.54 | -0.82 | -0.33 | -0.8 | -0.32 | 0 | -0.02 | 0.81 |
| 3 | 0.41 | 0.11 | -0.46 | -0.42 | -0.37 | -0.43 | -0.11 | 0.09 | -0.34 | 0.12 |
| 4 | 0.12 | 0.04 | -0.27 | -0.42 | -0.19 | -0.47 | -0.27 | 0.29 | -0.13 | 0.22 |
| 5 | 0.4 | -0.11 | -0.3 | -0.44 | -0.43 | -0.14 | -0.03 | -0.03 | -0.27 | 0.17 |
| 6 | 0.53 | -0.2 | -0.06 | -0.14 | -0.25 | -0.46 | 0.2 | 0.09 | -0.51 | 0.11 |
| 7 | 0 | -0.03 | 0.01 | -0.05 | -0.43 | -0.07 | 0.32 | 0.34 | -0.38 | 0.32 |
| 8 | 0.06 | -0.12 | -0.19 | 0.04 | 0.08 | 0.21 | 0.01 | 0.27 | -0.01 | 0.55 |

Mean $\log$ catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Mean Log q | -1.0887 | -0.765 | -0.5018 | -0.4048 | -0.3572 | -0.4057 | -0.6964 |
| S.E(Log q) | 0.5618 | 0.4264 | 0.3395 | 0.3723 | 0.3995 | 0.2815 | 0.3319 |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.

| Age | Slope | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 1.04 | -0.143 | 0.74 | 0.55 | 15 | 0.61 | -1.09 |
| 3 | 0.68 | 1.756 | 3.61 | 0.75 | 15 | 0.26 | -0.76 |
| 4 | 0.78 | 0.945 | 2.42 | 0.67 | 15 | 0.27 | -0.5 |
| 5 | 0.75 | 1.031 | 2.57 | 0.64 | 15 | 0.28 | -0.4 |
| 6 | 0.86 | 0.465 | 1.55 | 0.52 | 15 | 0.36 | -0.36 |
| 7 | 1.17 | -0.618 | -0.91 | 0.58 | 15 | 0.34 | -0.41 |
| 8 | 1.54 | -1.536 | -2.98 | 0.46 | 15 | 0.48 | -0.7 |

Terminal year survivor and F summaries :
Age 2 Catchability constant w.r.t. time and dependent on age
Year class $=2007$

| Fleet |  | Estimated | Int |  | Ext |  | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Survivors | s.e |  | s.e |  |  | Weights | F |
| Norw. Coast. survey | 5875 |  | 0.585 | 0 |  | 0 |  | 0.745 | 0 |
| F shrinkage mean | 250 |  | 1 |  |  |  |  | 0.255 | 0.011 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :--- | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 2623 | 0.51 | 1.6 | 2 | 3.158 | 0.001 |

Age 3 Catchability constant w.r.t. time and dependent on age
Year class $=2006$

| Fleet |  | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. survey | 5307 |  | 0.354 | 0.065 | 0.18 | 2 | 0.884 | 0.04 |
| F shrinkage mean | 2927 |  | 1 |  |  |  | 0.116 | 0.071 |
| Weighted prediction: |  |  |  |  |  |  |  |  |
| Survivors | Int |  | Ext | N | Var | F |  |  |
| at end of year | s.e |  | s.e |  | Ratio |  |  |  |
| 4953 | 0.33 |  | 0.15 | 3 | 0.449 | 0.043 |  |  |

Age 4 Catchability constant w.r.t. time and dependent on age
Year class $=2005$

| Fleet |  | Estimated Survivors | Int <br> s.e | Ext <br> s.e | Var <br> Ratio | N | Scaled Weights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Norw. Coast. survey | 4907 |  | 0.25 | 0.176 | 0.7 | 3 | 0.928 | 0.177 |
| F shrinkage mean | 4094 |  | 1 |  |  |  | 0.072 | 0.209 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors | Int |  | Ext | N | Var | F |  |  |
| at end of year | s.e |  | s.e |  | Ratio |  |  |  |
| 4843 | 0.24 |  | 0.14 | 4 | 0.58 | 0.18 |  |  |

Age 5 Catchability constant w.r.t. time and dependent on age
Year class $=2004$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Survivors | s.e | s.e | Ratio |  | Weights | F |  |
| Norw. Coast. survey | 2876 |  | 0.212 | 0.098 | 0.46 | 4 | 0.931 | 0.338 |  |
| F shrinkage mean | 2995 |  | 1 |  |  |  |  | 0.069 | 0.327 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :--- | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 2885 | 0.21 | 0.08 | 5 | 0.391 | 0.338 |

1
Age 6 Catchability constant w.r.t. time and dependent on age
Year class $=2003$

| Fleet |  | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. survey | 1693 |  | 0.193 | 0.153 | 0.79 | 5 | 0.921 | 0.555 |
| F shrinkage mean | 3035 |  | 1 |  |  |  | 0.079 | 0.346 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors | Int |  | Ext | N | Var | F |  |  |
| at end of year | s.e |  | s.e |  | Ratio |  |  |  |
| 1774 | 0.19 |  | 0.15 | 6 | 0.772 | 0.536 |  |  |

Age 7 Catchability constant w.r.t. time and dependent on age
Year class $=2002$

| Fleet |  | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. survey | 1132 |  | 0.171 | 0.153 | 0.89 | 6 | 0.932 | 0.616 |
| F shrinkage mean | 1834 |  | 1 |  |  |  | 0.068 | 0.422 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors | Int |  | Ext | N | Var | F |  |  |
| at end of year | s.e |  | s.e |  | Ratio |  |  |  |
| 1169 | 0.17 |  | 0.14 | 7 | 0.832 | 0.602 |  |  |

Age 8 Catchability constant w.r.t. time and dependent on age
Year class $=2001$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. survey | 786 |  | 0.16 | 0.178 | 1.11 | 7 | 0.941 | 0.487 |


| F shrinkage mean | 1106 | 1 | 0.059 | 0.369 |
| :--- | :--- | :--- | :--- | :--- |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :--- | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 802 | 0.16 | 0.16 | 8 | 1.008 | 0.481 |

Age 9 Catchability constant w.r.t. time and age (fixed at the value for age) 8
Year class $=2000$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :---: | :---: |
|  |  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. survey | 454 |  | 0.166 | 0.097 | 0.59 | 7 | 0.918 | 0.295 |
| F shrinkage mean | 244 |  | 1 |  |  |  | 0.082 | 0.493 |

Weighted prediction :

| Survivors | Int | Ext | N | Var | F |
| :--- | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 432 | 0.17 | 0.11 | 8 | 0.631 | 0.309 |

Table 2.16. Norwegian Coastal Cod. Summary output from XSA run for 2009 updated.

Terminal Fs derived using XSA (With F shrinkage)

|  | RECRUITS | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | SOPCOFAC | FBAR 4-7 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Age 2 |  |  |  |  |  |  |
| 1984 | 87915 | 310202 | 140813 | 74824 | 0.5314 | 1.0002 | 0.6221 |
| 1985 | 74401 | 293915 | 116933 | 75451 | 0.6452 | 1 | 0.5276 |
| 1986 | 35543 | 290530 | 122006 | 68905 | 0.5648 | 1.0001 | 0.5808 |
| 1987 | 36610 | 254620 | 114620 | 60972 | 0.5319 | 1.0001 | 0.4919 |
| 1988 | 39928 | 230358 | 117782 | 59294 | 0.5034 | 1.0001 | 0.6203 |
| 1989 | 43226 | 195737 | 93316 | 40285 | 0.4317 | 1 | 0.3763 |
| 1990 | 42297 | 209288 | 102157 | 28127 | 0.2753 | 1.0002 | 0.1845 |
| 1991 | 59811 | 244532 | 122385 | 24822 | 0.2028 | 1.0003 | 0.1714 |
| 1992 | 49295 | 286162 | 153360 | 41690 | 0.2718 | 1.0001 | 0.2356 |
| 1993 | 30294 | 298765 | 165785 | 52557 | 0.317 | 1 | 0.2375 |
| 1994 | 25179 | 298482 | 174761 | 54562 | 0.3122 | 1 | 0.2396 |
| 1995 | 33496 | 260753 | 161869 | 57207 | 0.3534 | 1.0001 | 0.3064 |
| 1996 | 39939 | 261957 | 173001 | 61776 | 0.3571 | 1.0001 | 0.383 |
| 1997 | 32577 | 204456 | 127641 | 63319 | 0.4961 | 1.0003 | 0.4071 |
| 1998 | 30422 | 175637 | 93903 | 51572 | 0.5492 | 0.9919 | 0.4502 |
| 1999 | 25013 | 154934 | 78153 | 40732 | 0.5212 | 1.0002 | 0.4557 |
| 2000 | 22754 | 138094 | 66319 | 36715 | 0.5536 | 0.9999 | 0.3879 |
| 2001 | 21125 | 129059 | 61969 | 29699 | 0.4793 | 1.0004 | 0.3299 |
| 2002 | 18070 | 154004 | 83131 | 40994 | 0.4931 | 1.0181 | 0.4054 |
| 2003 | 16965 | 105703 | 55358 | 34635 | 0.6257 | 1.0001 | 0.3893 |
| 2004 | 17204 | 108573 | 56881 | 24547 | 0.4315 | 0.9997 | 0.3435 |
| 2005 | 14265 | 100189 | 49732 | 22432 | 0.4511 | 1.0001 | 0.3102 |
| 2006 | 12136 | 112080 | 57270 | 26134 | 0.4563 | 0.9999 | 0.3564 |
| 2007 | 11555 | 108338 | 57439 | 23841 | 0.4151 | 0.9998 | 0.3161 |
| 2008 | 7808 | 99441 | 53352 | 25777 | 0.4831 | 0.9999 | 0.3322 |
| 2009 | 3206 | 81276 | 46617 | 24821 | 0.5324 | 1 | 0.4139 |

Arith.

| Mean 31963 | 196426 | 101790 | 44065 | .4533 | .3798 |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 Units | (Thousands) | (Tonnes) | (Tonnes) | (Tonnes) |  |  |

Table 2.17. Norwegian Coastal Cod. Diagnostic output from XSA run for 2010 with recreational and tourist fisheries included.

Lowestoft VPA Version 3.1
26/04/2010 15:31
Extended Survivors Analysis
Norwegian
Coastal Cod COMBSEX PLUSGROUP
CPUE data from file coast-9.txt
Catch data for 26 years. 1984 to 2009. Ages 2 to 10.

| Fleet |  | Last <br> year | First age | Last <br> age | Alpha | Beta |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Norw. Coast survey | 1995 | 2009 |  | 8 | 0.75 | 0.85 |  |  |  |  |
| Time series weights : <br> Tapered time weighting applied <br> Power = 3 over 20 years |  |  |  |  |  |  |  |  |  |  |
| Catchability analysis : <br> Catchability independent of stock size for all ages Catchability independent of age for ages $>=8$ |  |  |  |  |  |  |  |  |  |  |
| Terminal po Survivor of the fina S.E. of the Minimum estimates Prior weig | lation estima imates shrunk 2 years or the mean to which tandard error rived from ea ting not appli | towards th 4 oldest ag he estimat or populati h fleet = d | e mean F ges. <br> es are shrun ion <br> .300 | $\mathrm{nk}=1.000$ |  |  |  |  |  |  |
| Tuning had n Total absolut 29 and $30=$ Final year F | $t$ converged a residual betw .00475 lues | en 30 itera | ations |  |  |  |  |  |  |  |
| Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |  |  |
| Iteration 29 | 0.0008 | 0.0401 | 0.1885 | 0.355 | 0.5703 | 0.6075 | 0.4717 | 0.3113 |  |  |
| Iteration 30 | 0.0008 | 0.0401 | 0.1884 | 0.3548 | 0.5697 | 0.6066 | 0.47 | 0.3102 |  |  |
| Regression weights |  |  |  |  |  |  |  |  |  |  |
| Fishing mortalities |  |  |  |  |  |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 2 | 0.007 | 0.002 | 0.01 | 0.007 | 0.002 | 0.002 | 0.005 | 0.009 | 0.011 | 0.001 |
| 3 | 0.058 | 0.034 | 0.052 | 0.074 | 0.03 | 0.038 | 0.034 | 0.066 | 0.088 | 0.04 |
| 4 | 0.228 | 0.131 | 0.174 | 0.163 | 0.096 | 0.139 | 0.188 | 0.183 | 0.236 | 0.188 |
| 5 | 0.377 | 0.303 | 0.298 | 0.288 | 0.231 | 0.268 | 0.273 | 0.309 | 0.33 | 0.355 |
| 6 | 0.417 | 0.373 | 0.504 | 0.41 | 0.428 | 0.323 | 0.468 | 0.326 | 0.351 | 0.57 |
| 7 | 0.383 | 0.49 | 0.508 | 0.505 | 0.611 | 0.526 | 0.474 | 0.46 | 0.385 | 0.607 |
| 8 | 0.259 | 0.344 | 0.551 | 0.384 | 0.555 | 0.551 | 0.609 | 0.338 | 0.46 | 0.47 |
| 9 | 0.297 | 0.295 | 0.404 | 0.543 | 0.306 | 0.417 | 0.57 | 0.56 | 0.272 | 0.31 |
| XSA population numbers (Thousands) |  |  |  |  |  |  |  |  |  |  |
|  | AGE |  |  |  |  |  |  |  |  |  |
| YEAR | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |  |  |
| 2000 | $3.41 \mathrm{E}+04$ | $3.00 \mathrm{E}+04$ | $2.61 \mathrm{E}+04$ | $1.87 \mathrm{E}+04$ | $1.21 \mathrm{E}+04$ | $5.03 \mathrm{E}+03$ | $2.19 \mathrm{E}+03$ | $1.04 \mathrm{E}+03$ |  |  |
| 2001 | $3.19 \mathrm{E}+04$ | $2.77 \mathrm{E}+04$ | $2.32 \mathrm{E}+04$ | $1.70 \mathrm{E}+04$ | $1.05 \mathrm{E}+04$ | $6.51 \mathrm{E}+03$ | $2.81 \mathrm{E}+03$ | $1.39 \mathrm{E}+03$ |  |  |
| 2002 | $2.76 \mathrm{E}+04$ | $2.61 \mathrm{E}+04$ | $2.19 \mathrm{E}+04$ | $1.66 \mathrm{E}+04$ | $1.03 \mathrm{E}+04$ | $5.92 \mathrm{E}+03$ | $3.26 \mathrm{E}+03$ | $1.63 \mathrm{E}+03$ |  |  |
| 2003 | $2.58 \mathrm{E}+04$ | $2.24 \mathrm{E}+04$ | $2.03 \mathrm{E}+04$ | $1.51 \mathrm{E}+04$ | $1.01 \mathrm{E}+04$ | $5.08 \mathrm{E}+03$ | $2.92 \mathrm{E}+03$ | $1.54 \mathrm{E}+03$ |  |  |
| 2004 | $2.67 \mathrm{E}+04$ | $2.09 \mathrm{E}+04$ | $1.70 \mathrm{E}+04$ | $1.41 \mathrm{E}+04$ | $9.26 \mathrm{E}+03$ | $5.49 \mathrm{E}+03$ | $2.51 \mathrm{E}+03$ | $1.63 \mathrm{E}+03$ |  |  |
| 2005 | $2.21 \mathrm{E}+04$ | $2.18 \mathrm{E}+04$ | $1.66 \mathrm{E}+04$ | $1.26 \mathrm{E}+04$ | $9.17 \mathrm{E}+03$ | $4.94 \mathrm{E}+03$ | $2.44 \mathrm{E}+03$ | $1.18 \mathrm{E}+03$ |  |  |
| 2006 | $1.84 \mathrm{E}+04$ | $1.80 \mathrm{E}+04$ | $1.72 \mathrm{E}+04$ | $1.19 \mathrm{E}+04$ | $7.92 \mathrm{E}+03$ | $5.44 \mathrm{E}+03$ | $2.39 \mathrm{E}+03$ | $1.15 \mathrm{E}+03$ |  |  |
| 2007 | $1.72 \mathrm{E}+04$ | $1.50 \mathrm{E}+04$ | $1.43 \mathrm{E}+04$ | $1.17 \mathrm{E}+04$ | $7.38 \mathrm{E}+03$ | $4.06 \mathrm{E}+03$ | $2.77 \mathrm{E}+03$ | $1.06 \mathrm{E}+03$ |  |  |
| 2008 | $1.12 \mathrm{E}+04$ | $1.40 \mathrm{E}+04$ | $1.15 \mathrm{E}+04$ | $9.73 \mathrm{E}+03$ | $7.02 \mathrm{E}+03$ | $4.36 \mathrm{E}+03$ | $2.10 \mathrm{E}+03$ | $1.62 \mathrm{E}+03$ |  |  |
| 2009 | $3.91 \mathrm{E}+03$ | $9.06 \mathrm{E}+03$ | $1.05 \mathrm{E}+04$ | $7.43 \mathrm{E}+03$ | $5.73 \mathrm{E}+03$ | $4.05 \mathrm{E}+03$ | $2.43 \mathrm{E}+03$ | $1.09 \mathrm{E}+03$ |  |  |

Estimated population abundance at 1st Jan 2010
$0.00 \mathrm{E}+00 \quad 3.20 \mathrm{E}+03 \quad 7.13 \mathrm{E}+03 \quad 7.11 \mathrm{E}+03 \quad 4.27 \mathrm{E}+03 \quad 2.66 \mathrm{E}+03 \quad 1.81 \mathrm{E}+03 \quad 1.25 \mathrm{E}+03$
Taper weighted geometric mean of the VPA populations: $2.25 \mathrm{E}+04 \quad 2.21 \mathrm{E}+04 \quad 1.91 \mathrm{E}+04 \quad 1.43 \mathrm{E}+04 \quad 9.48 \mathrm{E}+03 \quad 5.61 \mathrm{E}+03 \quad 2.96 \mathrm{E}+03 \quad 1.56 \mathrm{E}+03$
Standard error of the weighted $\log (V P A$ populations $) ~: ~$

| 0.6908 | 0.4456 | 0.3708 | 0.3576 | 0.3349 | 0.3543 | 0.3776 | 0.3851 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Log catchability residuals.
Fleet : Norw. Coast. survey

| Age | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.84 | 0.45 | 0.74 | 0.52 | 0.46 |  |  |  |  |  |
| 3 | 0.68 | 0.93 | 0.93 | 0.51 | 0.35 |  |  |  |  |  |
| 4 | 0.69 | 0.67 | 0.76 | 0.32 | 0.21 |  |  |  |  |  |
| 5 | 0.47 | 0.94 | 1.03 | 0.37 | 0.18 |  |  |  |  |  |
| 6 | 0.12 | 0.12 | 1.39 | 0.21 | 0.13 |  |  |  |  |  |
| 7 | 0.06 | -0.28 | 0.46 | 0.3 | -0.27 |  |  |  |  |  |
| 8 | 0 | -0.25 | 0.18 | -0.87 | -0.46 |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 2 | 0.55 | 0.08 | -0.58 | -0.85 | -0.39 | -0.85 | -0.35 | -0.01 | 0.01 | 0.99 |
| 3 | 0.42 | 0.11 | -0.49 | -0.46 | -0.39 | -0.47 | -0.14 | 0.07 | -0.34 | 0.15 |
| 4 | 0.18 | 0.05 | -0.29 | -0.46 | -0.23 | -0.48 | -0.3 | 0.26 | -0.14 | 0.24 |
| 5 | 0.42 | -0.06 | -0.32 | -0.49 | -0.48 | -0.19 | -0.04 | -0.06 | -0.3 | 0.18 |
| 6 | 0.47 | -0.18 | -0.04 | -0.18 | -0.29 | -0.5 | 0.15 | 0.09 | -0.54 | 0.11 |
| 7 | 0 | -0.08 | 0.01 | -0.09 | -0.46 | -0.07 | 0.28 | 0.33 | -0.38 | 0.31 |
| 8 | 0.02 | -0.16 | -0.32 | 0.02 | 0.04 | 0.27 | 0.01 | 0.29 | 0.03 | 0.57 |

Mean log catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time

| Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Mean $\log q$ | -1.4723 | -1.1627 | -0.9001 | -0.8096 | -0.7698 | -0.8352 | -1.1554 |
| S.E(Log q) | 0.6103 | 0.4521 | 0.3729 | 0.4143 | 0.4176 | 0.2866 | 0.3371 |

Regression statistics :

Ages with q independent of year class strength and constant w.r.t. time.

| Age | Slope | t-value | Intercept | RSquare | No Pts | Reg s.e | Q |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 1.18 | -0.556 | -0.1 | 0.49 | 15 | 0.75 | -1.47 |
| 3 | 0.69 | 1.466 | 3.89 | 0.7 | 15 | 0.3 | -1.16 |
| 4 | 0.79 | 0.781 | 2.79 | 0.58 | 15 | 0.3 | -0.9 |
| 5 | 0.76 | 0.79 | 2.89 | 0.53 | 15 | 0.32 | -0.81 |
| 6 | 0.83 | 0.495 | 2.2 | 0.46 | 15 | 0.36 | -0.77 |
| 7 | 1.04 | -0.131 | 0.54 | 0.56 | 15 | 0.31 | -0.84 |
| 8 | 1.37 | -0.934 | -1.35 | 0.4 | 15 | 0.46 | -1.16 |

Terminal year survivor and $F$ summaries :
Age 2 Catchability constant w.r.t. time and dependent on age
Year class $=2007$

| Fleet | Estimated | Int |  | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e |  | s.e | Ratio |  | Weights | F |
| Norw. Coast. |  |  |  |  |  |  |  |  |
| survey | 8622 | 0.636 | 0 |  | 0 | 1 | 0.712 | 0 |
| F shrinkage |  |  |  |  |  |  |  |  |
| mean | 277 | 1 |  |  |  |  | 0.288 | 0.01 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors | Int | Ext |  | N | Var | F |  |  |
| at end of year | s.e | s.e |  |  | Ratio |  |  |  |
| 3203 | 0.54 | 1.85 | 2 |  | 3.439 | 0.001 |  |  |

Age 3 Catchability constant w.r.t. time and dependent on age
Year class $=2006$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. survey | 7882 | 0.379 | 0.07 | 0.18 | 2 | 0.87 | 0.036 |
| F shrinkage mean | 3638 | 1 |  |  |  | 0.13 | 0.077 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |
| at end of year | s.e | s.e |  | Ratio |  |  |  |
| 7127 | 0.35 | 0.2 | 3 | 0.572 | 0.04 |  |  |
| 1 |  |  |  |  |  |  |  |
| Age 4 Catchability constant w.r.t. time and dependent on ageYear class $=2005$ |  |  |  |  |  |  |  |
| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. survey | 7188 | 0.271 | 0.18 | 0.66 | 3 | 0.915 | 0.186 |
| F shrinkage mean | 6299 | 1 |  |  |  | 0.085 | 0.21 |
| Weighted pred | ction : |  |  |  |  |  |  |
| Survivors at end of year | s.e | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | N | Var <br> Ratio | F |  |  |
| 7108 | 0.26 | 0.14 | 4 | 0.543 | 0.188 |  |  |

Age 5 Catchability constant w.r.t. time and dependent on age Year class $=2004$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. survey | 4223 | 0.232 | 0.104 | 0.45 | 4 | 0.917 | 0.358 |
| F shrinkage mean | 4796 | 1 |  |  |  | 0.083 | 0.321 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |
| at end of year | s.e | s.e |  | Ratio |  |  |  |
| 4268 | 0.23 | 0.09 | 5 | 0.387 | 0.355 |  |  |

Age 6 Catchability constant w.r.t. time and dependent on age Year class $=2003$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. |  |  |  |  |  |  |  |
| survey | 2487 | 0.209 | 0.156 | 0.75 | 5 | 0.906 | 0.598 |
| F shrinkage |  |  |  |  |  |  |  |
| mean | 5018 | 1 |  |  |  | 0.094 | 0.341 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |
| at end of year | s.e | s.e |  | Ratio |  |  |  |
| 2656 | 0.21 | 0.16 | 6 | 0.776 | 0.57 |  |  |

Age 7 Catchability constant w.r.t. time and dependent on age

Year class $=2002$

| Fleet | Estimated | Int |  | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e |  | s.e | Ratio |  | Weights | F |
| Norw. Coast. |  |  |  |  |  |  |  |  |
| survey | 1746 | 0.18 | 0.16 |  | 0.89 | 6 | 0.927 | 0.622 |
| F shrinkage |  |  |  |  |  |  |  |  |
| mean | 2842 | 1 |  |  |  |  | 0.073 | 0.425 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N |  | Var | F |  |  |
| at end of year | s.e | s.e |  |  | Ratio |  |  |  |
| 1810 | 0.18 | 0.15 | 7 |  | 0.824 | 0.607 |  |  |

Age 8 Catchability constant w.r.t. time and dependent on age Year class $=2001$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. |  |  |  |  |  |  |  |
| F shrinkage mean | 1511 | 1 |  |  |  | 0.062 | 0.402 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |
| at end of year | s.e | s.e |  | Ratio |  |  |  |
| 1250 | 0.17 | 0.17 | 8 | 0.983 | 0.47 |  |  |

Age 9 Catchability constant w.r.t. time and age (fixed at the value for age) 8
Year class $=2000$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. |  |  |  |  |  |  |  |
| F shrinkage mean | 362 | 1 |  |  |  | 0.089 | 0.504 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |
| at end of year | s.e | s.e |  | Ratio |  |  |  |
| 654 | 0.18 | 0.11 | 8 | 0.627 | 0.31 |  |  |

Table 2.18. Norwegian Coastal Cod. Summary output from XSA run for 2010 with recreational and tourist fisheries included.

Terminal Fs derived using XSA (With F shrinkage)

| RECRUITS <br> Age 2 |  | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | SOPCOFAC | FBAR 4-7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |
| 1984 | 109445 | 363616 | 161708 | 88124 | 0.545 | 1.0001 | 0.6182 |
| 1985 | 98043 | 348434 | 134544 | 88851 | 0.6604 | 1 | 0.5235 |
| 1986 | 62696 | 351708 | 140662 | 82405 | 0.5858 | 1 | 0.5895 |
| 1987 | 49015 | 317004 | 133299 | 74472 | 0.5587 | 1 | 0.5067 |
| 1988 | 54425 | 294097 | 139624 | 72894 | 0.5221 | 1 | 0.6342 |
| 1989 | 62998 | 260422 | 117973 | 53985 | 0.4576 | 1 | 0.3834 |
| 1990 | 61657 | 279615 | 131106 | 42627 | 0.3251 | 1.0001 | 0.2371 |
| 1991 | 81509 | 325149 | 158032 | 40122 | 0.2539 | 1.0002 | 0.1941 |
| 1992 | 68237 | 371918 | 194241 | 57790 | 0.2975 | 1 | 0.249 |
| 1993 | 39633 | 385526 | 209475 | 67357 | 0.3216 | 1 | 0.2361 |
| 1994 | 33297 | 381398 | 218590 | 69262 | 0.3169 | 1.0001 | 0.2409 |
| 1995 | 45300 | 335804 | 206391 | 71907 | 0.3484 | 1.0001 | 0.3018 |
| 1996 | 57977 | 343119 | 224269 | 76276 | 0.3401 | 1 | 0.3616 |
| 1997 | 47218 | 274036 | 166813 | 77819 | 0.4665 | 1.0002 | 0.3962 |
| 1998 | 42205 | 246837 | 128856 | 66172 | 0.5135 | 0.9937 | 0.4108 |
| 1999 | 36936 | 222467 | 110823 | 54632 | 0.493 | 1.0001 | 0.413 |
| 2000 | 34100 | 202305 | 97923 | 50315 | 0.5138 | 0.9999 | 0.3512 |
| 2001 | 31920 | 190167 | 91132 | 43099 | 0.4729 | 1.0002 | 0.3244 |
| 2002 | 27591 | 222777 | 117369 | 54594 | 0.4651 | 1.0134 | 0.3708 |
| 2003 | 25759 | 160469 | 83672 | 48535 | 0.5801 | 1.0001 | 0.3416 |
| 2004 | 26727 | 165996 | 85725 | 37947 | 0.4427 | 0.9997 | 0.3415 |
| 2005 | 22064 | 150333 | 72584 | 35632 | 0.4909 | 1.0001 | 0.314 |
| 2006 | 18390 | 168916 | 84497 | 39134 | 0.4631 | 0.9998 | 0.3508 |
| 2007 | 17217 | 163168 | 85363 | 36841 | 0.4316 | 1 | 0.3192 |
| 2008 | 11181 | 145804 | 76271 | 38577 | 0.5058 | 0.9998 | 0.3254 |
| 2009 | 3914 | 119962 | 68660 | 37521 | 0.5465 | 1.0001 | 0.4299 |

Arith.

| Mean 44979 | 261194 | 132292 | 57957 | .4584 | .3756 |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 Units | (Thousands) | (Tonnes) | (Tonnes) | (Tonnes) |  |  |

## Table 2.19. Norwegian Coastal Cod. Diagnostic output from XSA run for 2010 with recreational and tourist fisheries included, and stock dependant catchabilities for ages 2 and 3.

Lowestoft VPA Version 3.1
26/04/2010 16:18
Extended Survivors Analysis

Norwegian Coastal Cod

## COMBSEX

CPUE data from file coast-9.txt
Catch data for 26 years. 1984 to 2009. Ages 2 to 10
Fleet Fi

Norw. Coast. survey
Time series weights :
Tapered time weighting applied
Power $=3$ over 20 years
Catchability analysis :
Catchability dependent on stock size for ages $<4$
Regression type $=C$
Minimum of 5 points used for regression
Survivor estimates shrunk to the population mean for ages $<4$
Catchability independent of age for ages $>=8$
Terminal population estimation :
Survivor estimates shrunk towards the mean F
of the final 2 years or the 4 oldest ages.
S.E. of the mean to which the estimates are shrunk $=1.000$

Minimum standard error for population
estimates derived from each fleet $=\quad .300$
Prior weighting not applied
Tuning had not converged after 30 iterations
Total absolute residual between iterations
29 and $30=.00808$
Final year $F$ values
Age 2
2
year
1995

PLUSGROUP

| Last | First | Last | Alpha | Beta |
| :---: | :---: | :---: | :---: | :---: |
| year | age | age |  |  |
| 2009 | 0 | 8 | 0.75 | 0.85 |

## 0

## 8

2009

4
5
6
$7 \quad 8$
9

| Iteration 29 |  | 0.0001 |  |  | 0.0181 |  | 0.1321 |  | 0.2646 |  | 0.4794 | 0.5589 | 0.412 | 0.2794 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Iteration 30 |  | 0.0001 |  |  | 0.018 |  | 0.1318 |  | 0.264 |  | 0.4781 | 0.5574 | 0.4092 | 0.2779 |  |  |
| Regression weights |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | 0.751 |  |  | 0.82 |  | 0.877 |  | 0.921 |  | 0.954 | 0.976 | 0.99 | 0.997 | 1 | 1 |
| Fishing mortalities |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |  |  |  |  |  |  |
| 2 | 0.007 | 0.001 | 0.01 | 0.007 | 0.002 | 0.002 | 0.004 | 0.007 | 0.005 | 0 |  |  |  |  |  |  |
| 3 | 0.057 | 0.034 | 0.051 | 0.073 | 0.029 | 0.037 | 0.032 | 0.054 | 0.0640 | 0.018 |  |  |  |  |  |  |
| 4 | 0.227 | 0.13 | 0.172 | 0.159 | 0.094 | 0.134 | 0.183 | 0.167 | 0.188 | 0.132 |  |  |  |  |  |  |
| 5 | 0.375 | 0.301 | 0.294 | 0.286 | 0.224 | 0.262 | 0.26 | 0.297 | 0.2940 | 0.264 |  |  |  |  |  |  |
| 6 | 0.422 | 0.371 | 0.5 | 0.402 | 0.422 | 0.31 | 0.453 | 0.305 | 0.3330 | 0.478 |  |  |  |  |  |  |
| 7 | 0.384 | 0.5 | 0.502 | 0.498 | 0.591 | 0.515 | 0.446 | 0.438 | 0.3510 | 0.557 |  |  |  |  |  |  |
| 8 | 0.251 | 0.346 | 0.569 | 0.377 | 0.541 | 0.52 | 0.586 | 0.309 | 0.4260 | 0.409 |  |  |  |  |  |  |
| 9 | 0.318 | 0.283 | 0.407 | 0.576 | 0.298 | 0.401 | 0.517 | 0.523 | 0.2420 | 0.278 |  |  |  |  |  |  |
| XSA population numbers (Thousands) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| YEAR | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |  |  |  |  |  |  |  |  |
| 2000 | $3.43 \mathrm{E}+04$ | $3.03 \mathrm{E}+04$ | $2.62 \mathrm{E}+04$ | $1.88 \mathrm{E}+04$ | $1.19 \mathrm{E}+04$ | $5.01 \mathrm{E}+03$ | $2.26 \mathrm{E}+03$ | $9.79 \mathrm{E}+02$ |  |  |  |  |  |  |  |  |
| 2001 | $3.26 \mathrm{E}+04$ | $2.79 \mathrm{E}+04$ | $2.34 \mathrm{E}+04$ | $1.71 \mathrm{E}+04$ | $1.06 \mathrm{E}+04$ | $6.41 \mathrm{E}+03$ | $2.79 \mathrm{E}+03$ | $1.44 \mathrm{E}+03$ |  |  |  |  |  |  |  |  |
| 2002 | $2.80 \mathrm{E}+04$ | $2.67 \mathrm{E}+04$ | $2.21 \mathrm{E}+04$ | $1.68 \mathrm{E}+04$ | $1.03 \mathrm{E}+04$ | $5.97 \mathrm{E}+03$ | $3.18 \mathrm{E}+03$ | $1.62 \mathrm{E}+03$ |  |  |  |  |  |  |  |  |
| 2003 | $2.67 \mathrm{E}+04$ | $2.27 \mathrm{E}+04$ | $2.08 \mathrm{E}+04$ | $1.52 \mathrm{E}+04$ | $1.03 \mathrm{E}+04$ | $5.14 \mathrm{E}+03$ | $2.96 \mathrm{E}+03$ | $1.47 \mathrm{E}+03$ |  |  |  |  |  |  |  |  |
| 2004 | $2.74 \mathrm{E}+04$ | $2.17 \mathrm{E}+04$ | $1.73 \mathrm{E}+04$ | $1.45 \mathrm{E}+04$ | $9.36 \mathrm{E}+03$ | $5.62 \mathrm{E}+03$ | $2.56 \mathrm{E}+03$ | $1.66 \mathrm{E}+03$ |  |  |  |  |  |  |  |  |
| 2005 | $2.39 \mathrm{E}+04$ | $2.24 \mathrm{E}+04$ | $1.73 \mathrm{E}+04$ | $1.29 \mathrm{E}+04$ | $9.50 \mathrm{E}+03$ | $5.02 \mathrm{E}+03$ | $2.55 \mathrm{E}+03$ | $1.22 \mathrm{E}+03$ |  |  |  |  |  |  |  |  |
| 2006 | $2.23 \mathrm{E}+04$ | $1.95 \mathrm{E}+04$ | $1.77 \mathrm{E}+04$ | $1.24 \mathrm{E}+04$ | $8.12 \mathrm{E}+03$ | $5.70 \mathrm{E}+03$ | $2.46 \mathrm{E}+03$ | $1.24 \mathrm{E}+03$ |  |  |  |  |  |  |  |  |
| 2007 | $2.33 \mathrm{E}+04$ | $1.82 \mathrm{E}+04$ | $1.55 \mathrm{E}+04$ | $1.21 \mathrm{E}+04$ | $7.81 \mathrm{E}+03$ | $4.22 \mathrm{E}+03$ | $2.99 \mathrm{E}+03$ | $1.12 \mathrm{E}+03$ |  |  |  |  |  |  |  |  |
| 2008 | $2.44 \mathrm{E}+04$ | $1.90 \mathrm{E}+04$ | $1.41 \mathrm{E}+04$ | $1.07 \mathrm{E}+04$ | $7.34 \mathrm{E}+03$ | $4.71 \mathrm{E}+03$ | $2.23 \mathrm{E}+03$ | $1.80 \mathrm{E}+03$ |  |  |  |  |  |  |  |  |
| 2009 | $2.25 \mathrm{E}+04$ | $1.99 \mathrm{E}+04$ | $1.46 \mathrm{E}+04$ | $9.56 \mathrm{E}+03$ | $6.55 \mathrm{E}+03$ | $4.31 \mathrm{E}+03$ | $2.72 \mathrm{E}+03$ | $1.19 \mathrm{E}+03$ |  |  |  |  |  |  |  |  |
| Estimated population abundance at 1st Jan 2010 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | $0.00 \mathrm{E}+00$ | $1.85 \mathrm{E}+04$ | $1.60 \mathrm{E}+04$ | $1.05 \mathrm{E}+04$ | $6.03 \mathrm{E}+03$ | $3.33 \mathrm{E}+03$ | $2.03 \mathrm{E}+03$ | $1.49 \mathrm{E}+03$ |  |  |  |  |  |  |  |  |
| Taper weighted geometric mean of the VPA populations: |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | $2.93 \mathrm{E}+04$ | $2.50 \mathrm{E}+04$ | $2.03 \mathrm{E}+04$ | $1.49 \mathrm{E}+04$ | $9.74 \mathrm{E}+03$ | $5.73 \mathrm{E}+03$ | $3.04 \mathrm{E}+03$ | $1.60 \mathrm{E}+03$ |  |  |  |  |  |  |  |  |
| Standard error of the weighted Log(VPA populations) : |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


|  | 0.2765 | 0.2875 | 0.2962 | 0.3053 | 0.309 | 0.3405 | 0.3642 | 0.3761 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Log catchability residuals. |  |  |  |  |  |  |  |  |  |  |
| Fleet : Norw. Coast. survey |  |  |  |  |  |  |  |  |  |  |
| Age | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |  |  |  |
| 2 | 0.09 | -0.21 | 0.02 | 0.02 | 0.08 |  |  |  |  |  |
| 3 | 0.25 | 0.17 | 0.03 | -0.01 | -0.01 |  |  |  |  |  |
| 4 | 0.77 | 0.74 | 0.84 | 0.4 | 0.28 |  |  |  |  |  |
| 5 | 0.53 | 1.01 | 1.07 | 0.43 | 0.24 |  |  |  |  |  |
| 6 | 0.14 | 0.18 | 1.45 | 0.23 | 0.17 |  |  |  |  |  |
| 7 | 0.1 | -0.27 | 0.51 | 0.36 | -0.26 |  |  |  |  |  |
| 8 | 0.02 | -0.21 | 0.18 | -0.78 | -0.37 |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 2 | 0.17 | 0.03 | -0.09 | -0.15 | -0.01 | -0.1 | 0.08 | 0.13 | -0.06 | 0 |
| 3 | 0.08 | 0.01 | -0.2 | -0.09 | -0.04 | -0.09 | 0.1 | 0.19 | -0.05 | -0.07 |
| 4 | 0.24 | 0.11 | -0.23 | -0.42 | -0.18 | -0.45 | -0.26 | 0.23 | -0.31 | -0.06 |
| 5 | 0.47 | -0.02 | -0.28 | -0.45 | -0.46 | -0.16 | -0.04 | -0.05 | -0.37 | -0.09 |
| 6 | 0.52 | -0.15 | -0.02 | -0.17 | -0.27 | -0.5 | 0.16 | 0.05 | -0.56 | -0.06 |
| 7 | 0.04 | -0.03 | 0.03 | -0.08 | -0.47 | -0.06 | 0.24 | 0.31 | -0.44 | 0.24 |
| 8 | 0.02 | -0.12 | -0.25 | 0.04 | 0.04 | 0.24 | 0 | 0.23 | -0.02 | 0.45 |
| Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time |  |  |  |  |  |  |  |  |  |  |
| Age | 4 | 5 | 6 | 7 | 8 |  |  |  |  |  |
| Mean Log q | -0.9725 | -0.863 | -0.8087 | -0.8689 | -1.1934 |  |  |  |  |  |
| S.E(Log q) | 0.3834 | 0.4277 | 0.4314 | 0.2915 | 0.2828 |  |  |  |  |  |
| Regression statistics: |  |  |  |  |  |  |  |  |  |  |
| Ages with q dependent on year class strength |  |  |  |  |  |  |  |  |  |  |
| Age | Slope | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean |  |  |  |
| 2 | 0.34 | 5.401 | 7.36 | 0.87 | 15 | 0.11 | -1.75 |  |  |  |
| 3 | 0.39 | 4.482 | 6.7 | 0.85 | 15 | 0.12 | -1.29 |  |  |  |
| Ages with q independent of year class strength and constant w.r.t. time. |  |  |  |  |  |  |  |  |  |  |
| Age | Slope | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean |  |  |  |
| 4 | 0.55 | 2.071 | 5 | 0.69 | 15 | 0.18 | -0.97 |  |  |  |


| 5 | 0.59 |  | 1.524 | 4.46 | 0.59 | 15 | 0.24 | -0.86 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | 0.75 |  | 0.72 | 2.9 | 0.46 | 15 | 0.33 | -0.81 |
| 7 | 0.98 |  | 0.058 | 1 | 0.56 | 15 | 0.3 | -0.87 |
| 8 | 1.17 |  | -0.554 | 0.06 | 0.53 | 15 | 0.34 | -1.19 |
| Terminal year survivor and F summaries : |  |  |  |  |  |  |  |  |
| Age 2 Catchability dependent on age and year class strength Year class $=2007$ |  |  |  |  |  |  |  |  |
| Fleet |  | Estimated Survivors | Int <br> s.e | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled <br> Weights | Estimated F |
| Norw. Coast. survey | 18430 |  | 0.3 | 0 | 0 | 1 | 0.459 | 0 |
| P shrinkage mean | 25017 |  | 0.29 |  |  |  | 0.5 | 0 |
| F shrinkage mean | 475 |  | 1 |  |  |  | 0.041 | 0.006 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors | Int |  | Ext | N | Var | F |  |  |
| at end of year | s.e |  | s.e |  | Ratio |  |  |  |
| 18460 | 0.2 |  | 0.55 | 3 | 2.693 | 0 |  |  |
| Age 3 Catchability dependent on age and year class strength |  |  |  |  |  |  |  |  |
| Year class $=2006$ |  |  |  |  |  |  |  |  |
| Fleet |  | Estimated | Int | Ext | Var | N | Scaled | Estimated |
|  |  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. survey | 14971 |  | 0.212 | 0.006 | 0.03 | 2 | 0.637 | 0.019 |
| P shrinkage mean | 20309 |  | 0.3 |  |  |  | 0.334 | 0.014 |
| F shrinkage mean | 4786 |  | 1 |  |  |  | 0.029 | 0.059 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors |  |  | Ext | N | Var | F |  |  |
| at end of year | s.e |  | s.e |  | Ratio |  |  |  |
| 16030 | 0.17 |  | 0.15 | 4 | 0.893 | 0.018 |  |  |
| Age 4 Catchability constant w.r.t. time and dependent on age |  |  |  |  |  |  |  |  |
| Year class $=2005$ |  |  |  |  |  |  |  |  |
| Fleet |  | Estimated | Int | Ext | Var | N | Scaled | Estimated |
|  |  | Survivors | s.e | s.e | Ratio |  | Weights |  |
| Norw. Coast. survey | 10618 |  | 0.188 | 0.063 | 0.33 | 3 | 0.959 | 0.13 |


| F shrinkage mean | 7557 |  | 1 |  |  |  | 0.041 | 0.178 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Weighted prediction: |  |  |  |  |  |  |  |  |
| Survivors | Int |  | Ext | N | Var | F |  |  |
| at end of year | s.e |  | s.e |  | Ratio |  |  |  |
| 10473 | 0.18 |  | 0.06 | 4 | 0.346 | 0.132 |  |  |
| Age 5 Catchability constant w.r.t. time and dependent on age |  |  |  |  |  |  |  |  |
| Year class $=2004$ |  |  |  |  |  |  |  |  |
| Fleet |  | Estimated | Int | Ext | Var | N | Scaled | Estimated |
|  |  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. survey | 6065 |  | 0.174 | 0.105 | 0.6 | 4 | 0.955 | 0.262 |
| F shrinkage mean | 5247 |  | 1 |  |  |  | 0.045 | 0.297 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors | Int |  | Ext | N | Var | F |  |  |
| at end of year | s.e |  | s.e |  | Ratio |  |  |  |
| 6025 | 0.17 |  | 0.09 | 5 | 0.522 | 0.264 |  |  |
| Age 6 Catchability constant w.r.t. time and dependent on age |  |  |  |  |  |  |  |  |
| Year class $=2003$ |  |  |  |  |  |  |  |  |
| Fleet |  | Estimated | Int | Ext | Var | N | Scaled | Estimated |
|  |  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. survey | 3236 |  | 0.165 | 0.091 | 0.55 | 5 | 0.942 | 0.488 |
| F shrinkage mean | 5379 |  | 1 |  |  |  | 0.058 | 0.321 |
| Weighted prediction: |  |  |  |  |  |  |  |  |
| Survivors | Int |  | Ext | N | Var | F |  |  |
| at end of year | s.e |  | s.e |  | Ratio |  |  |  |
| 3333 | 0.17 |  | 0.1 | 6 | 0.581 | 0.478 |  |  |
| Age 7 Catchability constant w.r.t. time and dependent on age |  |  |  |  |  |  |  |  |
| Year class $=2002$ |  |  |  |  |  |  |  |  |
| Fleet |  | Estimated | Int | Ext | Var | N | Scaled | Estimated |
|  |  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. survey | 1975 |  | 0.154 | 0.117 | 0.76 | 6 | 0.943 | 0.567 |
| F shrinkage mean | 3091 |  | 1 |  |  |  | 0.057 | 0.397 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors | Int |  | Ext | N | Var | F |  |  |


| at end of year | s.e |  | s.e |  | Ratio |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2026 | 0.16 |  | 0.11 | 7 | 0.72 | 0.557 |  |  |
| Age 8 Catchability constant w.r.t. time and dependent on age Year class $=2001$ |  |  |  |  |  |  |  |  |
| Fleet |  | Estimated Survivors | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled <br> Weights | Estimated F |
| Norw. Coast. survey | 1478 |  | 0.144 | 0.148 | 1.03 | 7 | 0.956 | 0.409 |
| F shrinkage mean | 1668 |  | 1 |  |  |  | 0.044 | 0.37 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors |  |  | Ext | N | Var | F |  |  |
| at end of year | s.e |  | s.e |  | Ratio |  |  |  |
| 1486 | 0.14 |  | 0.13 | 8 | 0.933 | 0.409 |  |  |
| Age 9 Catchability constant w.r.t. time and age (fixed at the value for age) |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |
| Fleet |  | Estimated | Int | Ext | Var | N | Scaled | Estimated |
|  |  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| Norw. Coast. survey | 773 |  | 0.151 | 0.067 | 0.45 | 7 | 0.934 | 0.268 |
| F shrinkage mean | 441 |  | 1 |  |  |  | 0.066 | 0.43 |
| Weighted prediction : |  |  |  |  |  |  |  |  |
| Survivors |  |  | Ext | N | Var | F |  |  |
| at end of year | s.e |  | s.e |  | Ratio |  |  |  |
| 745 | 0.16 |  | 0.08 | 8 | 0.521 | 0.278 |  |  |

Table 2.20. Norwegian Coastal Cod. Summary output from XSA run for 2010 with recreational and tourist fisheries included, and stock dependant catchabilities for age 2 and 3.


Arith.

| Mean | 46800 | 265338 | 133518 | 57957 | .4515 | .3694 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 Units | (Thousands) | (Tonnes) | (Tonnes) | (Tonnes) |  |  |

Table 2.21. Norwegian Coastal Cod. Output from final standard VPA run for 2010.

Table 8 Fishing mortality (F) at age

| YEAR | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |
|  | 0.0152 | 0.0411 | 0.0871 | 0.0063 | 0.0153 | 0.0081 |
|  | 0.0945 | 0.1409 | 0.1082 | 0.071 | 0.097 | 0.0515 |
|  | 0.2393 | 0.2555 | 0.3157 | 0.2732 | 0.1641 | 0.1154 |
|  | 0.3664 | 0.476 | 0.4854 | 0.5628 | 0.2952 | 0.1984 |
|  | 0.6625 | 0.6259 | 0.6725 | 0.457 | 0.7948 | 0.4143 |
|  | 1.2029 | 0.7425 | 0.8887 | 0.7407 | 1.2752 | 0.8007 |
|  | 1.0048 | 0.565 | 0.9661 | 0.7009 | 1.1625 | 1.0172 |
|  | 0.8193 | 0.6062 | 0.7607 | 0.6187 | 0.8948 | 0.6159 |
| +gp | 0.8193 | 0.6062 | 0.7607 | 0.6187 | 0.8948 | 0.6159 |
| FBAR 4-7 | 0.6178 | 0.525 | 0.5906 | 0.5084 | 0.6323 | 0.3822 |


| YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| AGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.0074 | 0.0179 | 0.0077 | 0.0014 | 0.013 | 0.0203 | 0.0237 | 0.0328 | 0.0213 | 0.009 |
|  | 0.0407 | 0.0762 | 0.0333 | 0.0338 | 0.045 | 0.0517 | 0.0951 | 0.1051 | 0.1243 | 0.0603 |
|  | 0.0842 | 0.1304 | 0.1754 | 0.0658 | 0.0912 | 0.1492 | 0.1712 | 0.1876 | 0.2634 | 0.1682 |
|  | 0.1088 | 0.2132 | 0.2558 | 0.1442 | 0.1496 | 0.2562 | 0.4227 | 0.2615 | 0.3878 | 0.3861 |
|  | 0.2919 | 0.2067 | 0.2688 | 0.239 | 0.2548 | 0.3277 | 0.3995 | 0.4672 | 0.4221 | 0.5099 |
|  | 0.4623 | 0.2269 | 0.2968 | 0.4951 | 0.4669 | 0.4794 | 0.4567 | 0.6889 | 0.5805 | 0.5776 |
|  | 0.5 | 0.1989 | 0.3397 | 0.4384 | 0.5041 | 0.4141 | 0.636 | 0.7114 | 0.7115 | 0.6273 |
|  | 0.3436 | 0.212 | 0.2916 | 0.3312 | 0.3461 | 0.3712 | 0.4732 | 0.6271 | 0.4956 | 0.8755 |
| +gp | 0.3436 | 0.212 | 0.2916 | 0.3312 | 0.3461 | 0.3712 | 0.4732 | 0.6271 | 0.4956 | 0.8755 |
| FBAR 4-7 | 0.2368 | 0.1943 | 0.2492 | 0.236 | 0.2406 | 0.3031 | 0.3625 | 0.4013 | 0.4135 | 0.4104 |


|  | YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | FBAR $^{* *}$ _** |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

## AGE

    \(\begin{array}{llllllllllll}\text { +gp } & 0.3175 & 0.2834 & 0.4074 & 0.5759 & 0.2984 & 0.4008 & 0.5166 & 0.5234 & 0.2416 & 0.4206\end{array}\)
    0 FBAR 4-7 0.3526

Table 2.21. Norwegian Coastal Cod. Output from final standard VPA run for 2010 continued.

|  | Table 10 | Stock number at age (start of year) |  |  |  |  | Numbers*10**-3 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |  |  |  |  |  |  |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |
| 2 |  | 108384 | 97422 | 62363 | 48753 | 54062 | 62736 |  |  |  |  |  |  |
| 3 |  | 63659 | 87401 | 76550 | 46799 | 39664 | 43591 |  |  |  |  |  |  |
| 4 |  | 46837 | 47422 | 62151 | 56247 | 35689 | 29473 |  |  |  |  |  |  |
| 5 |  | 31977 | 30186 | 30071 | 37109 | 35041 | 24796 |  |  |  |  |  |  |
| 6 |  | 16215 | 18150 | 15354 | 15153 | 17306 | 21355 |  |  |  |  |  |  |
| 7 |  | 8600 | 6844 | 7947 | 6417 | 7855 | 6400 |  |  |  |  |  |  |
| 8 |  | 4104 | 2115 | 2667 | 2675 | 2505 | 1797 |  |  |  |  |  |  |
| 9 |  | 1853 | 1230 | 984 | 831 | 1087 | 641 |  |  |  |  |  |  |
|  | +gp | 1215 | 634 | 833 | 353 | 612 | 204 |  |  |  |  |  |  |
| 0 | TOTAL | 282845 | 291404 | 258920 | 214336 | 193821 | 190992 |  |  |  |  |  |  |
|  | YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |
| 2 |  | 61090 | 81382 | 67144 | 39068 | 33267 | 44916 | 57255 | 47031 | 42033 | 36994 |  |  |
| 3 |  | 50949 | 49648 | 65448 | 54549 | 31940 | 26885 | 36036 | 45780 | 37262 | 33689 |  |  |
| 4 |  | 33898 | 40050 | 37666 | 51828 | 43176 | 25000 | 20903 | 26828 | 33744 | 26942 |  |  |
| 5 |  | 21500 | 25512 | 28781 | 25877 | 39731 | 32268 | 17631 | 14422 | 18207 | 21230 |  |  |
| 6 |  | 16649 | 15789 | 16876 | 18245 | 18342 | 28009 | 20447 | 9459 | 9090 | 10114 |  |  |
| 7 |  | 11554 | 10181 | 10513 | 10560 | 11762 | 11639 | 16525 | 11228 | 4854 | 4880 |  |  |
| 8 |  | 2353 | 5958 | 6643 | 6397 | 5270 | 6037 | 5900 | 8569 | 4616 | 2224 |  |  |
| 9 |  | 532 | 1168 | 3998 | 3872 | 3379 | 2606 | 3267 | 2557 | 3444 | 1855 |  |  |
|  | +gp | 275 | 541 | 1211 | 2685 | 2992 | 1531 | 1577 | 1603 | 960 | 597 |  |  |
| 0 | TOTAL | 198800 | 230228 | 238282 | 213082 | 189858 | 178893 | 179541 | 167477 | 154211 | 138524 |  |  |
|  | YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | GMST |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |
| 2 |  | 34051 | 32448 | 26396 | 25708 | 29632 | 27864 | 21374 | 18883 | 24436 | 22068 | 0 | 43594 |
| 3 |  | 30016 | 27681 | 26526 | 21387 | 20898 | 24226 | 22780 | 17419 | 15336 | 19910 | 18065 | 37294 |
| 4 |  | 25968 | 23199 | 21897 | 20643 | 16203 | 16604 | 19163 | 18153 | 13484 | 11595 | 16010 | 30090 |
| 5 |  | 18643 | 16926 | 16675 | 15081 | 14416 | 11997 | 11830 | 13282 | 12921 | 9076 | 8027 | 21579 |
| 6 |  | 11815 | 10486 | 10245 | 10168 | 9271 | 9430 | 7407 | 7382 | 8352 | 8353 | 5626 | 13415 |
| 7 |  | 4973 | 6341 | 5926 | 5090 | 5564 | 4972 | 5663 | 3666 | 4374 | 5146 | 4818 | 7364 |
| 8 |  | 2242 | 2772 | 3147 | 2940 | 2533 | 2524 | 2431 | 2971 | 1788 | 2449 | 2720 | 3435 |
| 9 |  | 972 | 1428 | 1605 | 1458 | 1651 | 1208 | 1228 | 1109 | 1787 | 836 | 1265 | 1572 |
|  | +gp | 613 | 530 | 1235 | 247 | 1123 | 475 | 607 | 526 | 870 | 881 | 923 |  |
| 0 | TOTAL | 129294 | 121810 | 113652 | 102721 | 101291 | 99299 | 92486 | 83390 | 83348 | 80314 | 57454 |  |

Table 2.21. Norwegian Coastal Cod. Output from final standard VPA run for 2010 continued.

|  | Table 12 S | Stock biom | ass at a | (start | year) |  | onnes |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |  |  |  |  |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
| 2 |  | 34791 | 31272 | 20019 | 15650 | 17354 | 20138 |  |  |  |  |
| 3 |  | 48254 | 66250 | 58025 | 35474 | 30065 | 33042 |  |  |  |  |
| 4 |  | 69272 | 70137 | 91922 | 83189 | 52784 | 43590 |  |  |  |  |
| 5 |  | 68336 | 64508 | 64261 | 79303 | 74882 | 52989 |  |  |  |  |
| 6 |  | 45629 | 51073 | 43207 | 42639 | 48700 | 60092 |  |  |  |  |
| 7 |  | 40610 | 32318 | 37525 | 30299 | 37092 | 30219 |  |  |  |  |
| 8 |  | 27438 | 14136 | 17827 | 17885 | 16745 | 12011 |  |  |  |  |
| 9 |  | 12934 | 8588 | 6868 | 5800 | 7585 | 4476 |  |  |  |  |
|  | +gp | 11810 | 6168 | 8095 | 3429 | 5954 | 1986 |  |  |  |  |
| 0 | TOTALBIO | 359074 | 344451 | 347749 | 313668 | 291160 | 258544 |  |  |  |  |
|  | YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
| 2 |  | 19610 | 26123 | 21553 | 12541 | 10679 | 13385 | 15459 | 10911 | 13577 | 11764 |
| 3 |  | 38620 | 37633 | 49610 | 41348 | 24211 | 18820 | 25838 | 30993 | 31077 | 27086 |
| 4 |  | 50136 | 59234 | 55709 | 76653 | 63857 | 33450 | 29996 | 36566 | 46094 | 42002 |
| 5 |  | 45946 | 54519 | 61505 | 55298 | 84906 | 63666 | 36038 | 27445 | 37779 | 43351 |
| 6 |  | 46849 | 44430 | 47490 | 51342 | 51614 | 74195 | 55085 | 26637 | 27389 | 28300 |
| 7 |  | 54556 | 48072 | 49644 | 49864 | 55539 | 48467 | 79600 | 43035 | 20653 | 22828 |
| 8 |  | 15728 | 39827 | 44410 | 42765 | 35229 | 42567 | 37052 | 50119 | 24487 | 15903 |
| 9 |  | 3713 | 8155 | 27907 | 27030 | 23582 | 16713 | 37127 | 24550 | 28760 | 16620 |
|  | +gp | 2678 | 5260 | 11776 | 26110 | 29094 | 21936 | 24708 | 20902 | 17304 | 10942 |
| 0 | TOTALBIO | 277835 | 323254 | 369604 | 382952 | 378710 | 333198 | 340903 | 271159 | 247120 | 218796 |
|  | YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
| 2 |  | 11782 | 11259 | 11350 | 7918 | 10045 | 11341 | 10473 | 9781 | 12413 | 9578 |
| 3 |  | 23323 | 24304 | 23343 | 14671 | 17429 | 20495 | 25628 | 20641 | 18526 | 22219 |
| 4 |  | 37862 | 35796 | 37182 | 26815 | 26152 | 29024 | 34723 | 36507 | 28249 | 23225 |
| 5 |  | 42804 | 37458 | 40888 | 32409 | 32710 | 26393 | 30274 | 33206 | 38594 | 26266 |
| 6 |  | 32313 | 30011 | 36246 | 31875 | 30500 | 25394 | 26511 | 23327 | 30660 | 30339 |
| 7 |  | 20131 | 21057 | 26056 | 20604 | 22947 | 18977 | 22449 | 15545 | 17393 | 25086 |
| 8 |  | 15721 | 13441 | 13189 | 14723 | 11951 | 9584 | 11725 | 20218 | 7842 | 13223 |
| 9 |  | 8969 | 10483 | 11307 | 8439 | 8216 | 6456 | 9006 | 12251 | 9676 | 5121 |
|  | +gp | 7529 | 6112 | 19285 | 2488 | 7143 | 7038 | 8898 | 7855 | 10060 | 4156 |
| 0 | TOTALBIO | 200434 | 189920 | 218845 | 159943 | 167093 | 154703 | 179687 | 179332 | 173413 | 159214 |

Table 2.21. Norwegian Coastal Cod. Output from final standard VPA run for 2010 continued.

Table 13 Spawning stock biomass at age (spawning time) Tonnes

| YEAR | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

AGE

| 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 965 | 1325 | 1161 | 709 | 601 | 661 |
| 4 | 11084 | 11222 | 14707 | 13310 | 8445 | 6974 |
| 5 | 31435 | 29674 | 29560 | 36479 | 34446 | 24375 |
| 6 | 31484 | 35240 | 29813 | 29421 | 33603 | 41463 |
| 7 | 35330 | 28117 | 32647 | 26360 | 32270 | 26290 |
| 8 | 24969 | 12863 | 16223 | 16275 | 15238 | 10930 |
| 9 | 12417 | 8244 | 6593 | 5568 | 7282 | 4297 |
|  | +gp | 11810 | 6168 | 8095 | 3429 | 5954 |
| 0 | TOTSPBIO | 159493 | 132854 | 138799 | 131553 | 137838 |

Table 13 Spawning stock biomass at age (spawning time) Tonnes
$\begin{array}{lllllllllllll}\text { YEAR } & 1990 & 1991 & 1992 & 1993 & 1994 & 1995 & 1996 & 1997 & 1998 & 1999\end{array}$

AGE

| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 772 | 753 | 992 | 827 | 484 | 376 | 517 | 620 | 622 | 542 |
| 4 | 8022 | 9477 | 8913 | 12265 | 10217 | 5352 | 4799 | 5851 | 7375 | 6720 |
| 5 | 21135 | 25079 | 28292 | 25437 | 39057 | 29286 | 16577 | 12625 | 17378 | 19942 |
| 6 | 32326 | 30657 | 32768 | 35426 | 35614 | 51194 | 38009 | 18380 | 18899 | 19527 |
| 7 | 47464 | 41823 | 43190 | 43382 | 48319 | 42166 | 69252 | 37441 | 17968 | 19860 |
| 8 | 14312 | 36243 | 40413 | 38916 | 32058 | 38736 | 33717 | 45608 | 22283 | 14472 |
| 9 | 3565 | 7829 | 26791 | 25949 | 22639 | 16045 | 35642 | 23568 | 27610 | 15955 |
|  |  | 2678 | 5260 | 11776 | 26110 | 29094 | 21936 | 24708 | 20902 | 17304 |
| 10942 |  |  |  |  |  |  |  |  |  |  |
| 0 | TOTSPBIO | 130274 | 157120 | 193136 | 208312 | 217481 | 205091 | 223221 | 164994 | 129439 |
| 107960 |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  | YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| 2009 |  |  |  |  |  |  |  |  |  |  |

AGE

| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 466 | 486 | 467 | 293 | 349 | 410 | 513 | 413 | 371 | 444 |
| 4 | 6058 | 5727 | 5949 | 4290 | 4184 | 4644 | 5556 | 5841 | 4520 | 3716 |
| 5 | 19690 | 17230 | 18808 | 14908 | 15047 | 12141 | 13926 | 15275 | 17753 | 12082 |
| 6 | 22296 | 20707 | 25010 | 21994 | 21045 | 17522 | 18292 | 16096 | 21155 | 20934 |
| 7 | 17514 | 18320 | 22669 | 17925 | 19964 | 16510 | 19531 | 13525 | 15132 | 21825 |
| 8 | 14307 | 12231 | 12002 | 13398 | 10875 | 8721 | 10669 | 18399 | 7136 | 12033 |
| 9 | 8610 | 10064 | 10855 | 8102 | 7887 | 6198 | 8646 | 11761 | 9289 | 4916 |
|  | +gp | 7529 | 6112 | 19285 | 2488 | 7143 | 7038 | 8898 | 7855 | 10060 |
|  | TOTSPBIO | 96470 | 90878 | 115044 | 83399 | 86494 | 73185 | 86030 | 89164 | 85416 |

Table 2.22. Norwegian Coastal Cod. Summary utput from final standard VPA run for 2010 continued.

Traditional vpa using file input for terminal F

|  | RECRUITS |  | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | FBAR $4-7$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Age 2 |  |  |  |  |  |  |
| 1984 | 108384 | 359074 | 159493 | 88124 | 0.5525 | 0.6178 |  |
| 1985 | 97422 | 344451 | 132854 | 88851 | 0.6688 | 0.525 |  |
| 1986 | 62363 | 347749 | 138799 | 82405 | 0.5937 | 0.5906 |  |
| 1987 | 48753 | 313668 | 131553 | 74472 | 0.5661 | 0.5084 |  |
| 1988 | 54062 | 291160 | 137838 | 72894 | 0.5288 | 0.6323 |  |
| 1989 | 62736 | 258544 | 116978 | 53985 | 0.4615 | 0.3822 |  |
| 1990 | 61090 | 277835 | 130274 | 42627 | 0.3272 | 0.2368 |  |
| 1991 | 81382 | 323254 | 157120 | 40122 | 0.2554 | 0.1943 |  |
| 1992 | 67144 | 369604 | 193136 | 57790 | 0.2992 | 0.2492 |  |
| 1993 | 39068 | 382952 | 208312 | 67357 | 0.3233 | 0.236 |  |
| 1994 | 33267 | 378710 | 217481 | 69262 | 0.3185 | 0.2406 |  |
| 1995 | 44916 | 333198 | 205091 | 71907 | 0.3506 | 0.3031 |  |
| 1996 | 57255 | 340903 | 223221 | 76276 | 0.3417 | 0.3625 |  |
| 1997 | 47031 | 271159 | 164994 | 77819 | 0.4716 | 0.4013 |  |
| 1998 | 42033 | 247120 | 129439 | 66172 | 0.5112 | 0.4135 |  |
| 1999 | 36994 | 218796 | 107960 | 54632 | 0.506 | 0.4104 |  |
| 2000 | 34051 | 200434 | 96470 | 50315 | 0.5216 | 0.3526 |  |
| 2001 | 32448 | 189920 | 90878 | 43099 | 0.4743 | 0.3259 |  |
| 2002 | 26396 | 218845 | 115044 | 54594 | 0.4745 | 0.367 |  |
| 2003 | 25708 | 159943 | 83399 | 48535 | 0.582 | 0.3366 |  |
| 2004 | 29632 | 167093 | 86494 | 37947 | 0.4387 | 0.3346 |  |
| 2005 | 27864 | 154703 | 73185 | 35632 | 0.4869 | 0.3116 |  |
| 2006 | 21374 | 179687 | 86030 | 39134 | 0.4549 | 0.3467 |  |
| 2007 | 18883 | 179332 | 89164 | 36841 | 0.4132 | 0.3113 |  |
| 2008 | 24436 | 173413 | 85416 | 38577 | 0.4516 | 0.2741 |  |
| 2009 | 22068 | 159214 | 80107 | 37521 | 0.4684 | 0.3085 |  |

Arith.

| Mean | 46414 | 263106 | 132336 | 57957 | 0.4555 | 0.3682 |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 Units | (Thousands) | (Tonnes) | (Tonnes) | (Tonnes) |  |  |

Table 2.23. Norwegian Coastal Cod. Input to the predictions using fixed fishing pattern.

MFDP version 1a
Run: Pred2
Time and date: 20:23 26.04.2010
Fbar age range: 4-7
2010

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 21500 | 0.2 | 0 | 0 | 0 | 0.4867 | 0.0045 | 1.1457 |
| 3 | 18065 | 0.2 | 0.02 | 0 | 0 | 1.1697 | 0.0530 | 1.6987 |
| 4 | 16010 | 0.2 | 0.16 | 0 | 0 | 2.0363 | 0.1738 | 2.3117 |
| 5 | 8027 | 0.2 | 0.46 | 0 | 0 | 2.7937 | 0.2686 | 3.2653 |
| 6 | 5626 | 0.2 | 0.69 | 0 | 0 | 3.4877 | 0.3306 | 4.1713 |
| 7 | 4818 | 0.2 | 0.87 | 0 | 0 | 4.3640 | 0.4610 | 4.8990 |
| 8 | 2720 | 0.2 | 0.91 | 0 | 0 | 5.5310 | 0.4584 | 5.9127 |
| 9 | 1265 | 0.2 | 0.96 | 0 | 0 | 7.5303 | 0.4091 | 6.6030 |
| 10 | 923 | 0.2 | 1 | 0 | 0 | 10.4027 | 0.4091 | 8.2993 |

2011

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 21500 | 0.2 | 0 | 0 | 0 | 0.4867 | 0.0045 | 1.1457 |
| 3 | $\cdot$ | 0.2 | 0.02 | 0 | 0 | 1.1697 | 0.0530 | 1.6987 |
| 4 | $\cdot$ | 0.2 | 0.16 | 0 | 0 | 2.0363 | 0.1738 | 2.3117 |
| 5 | $\cdot$ | 0.2 | 0.46 | 0 | 0 | 2.7937 | 0.2686 | 3.2653 |
| 6 | $\cdot$ | 0.2 | 0.69 | 0 | 0 | 3.4877 | 0.3306 | 4.1713 |
| 7 | $\cdot$ | 0.2 | 0.87 | 0 | 0 | 4.3640 | 0.4610 | 4.8990 |
| 8 | $\cdot$ | 0.2 | 0.91 | 0 | 0 | 5.5310 | 0.4584 | 5.9127 |
| 9 | $\cdot$ | 0.2 | 0.96 | 0 | 0 | 7.5303 | 0.4091 | 6.6030 |
| 10 | . | 0.2 | 1 | 0 | 0 | 10.4027 | 0.4091 | 8.2993 |


| 2012 |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 21500 | 0.2 | 0 | 0 | 0 | 0.4867 | 0.0045 | 1.1457 |
| 3 | $\cdot$ | 0.2 | 0.02 | 0 | 0 | 1.1697 | 0.0530 | 1.6987 |
| 4 | $\cdot$ | 0.2 | 0.16 | 0 | 0 | 2.0363 | 0.1738 | 2.3117 |
| 5 | $\cdot$ | 0.2 | 0.46 | 0 | 0 | 2.7937 | 0.2686 | 3.2653 |
| 6 | $\cdot$ | 0.2 | 0.69 | 0 | 0 | 3.4877 | 0.3306 | 4.1713 |
| 7 | $\cdot$ | 0.2 | 0.87 | 0 | 0 | 4.3640 | 0.4610 | 4.8990 |
| 8 | $\cdot$ | 0.2 | 0.91 | 0 | 0 | 5.5310 | 0.4584 | 5.9127 |
| 9 | $\cdot$ | 0.2 | 0.96 | 0 | 0 | 7.5303 | 0.4091 | 6.6030 |
| 10 | . | 0.2 | 1 | 0 | 0 | 10.4027 | 0.4091 | 8.2993 |


| 2013 |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 21500 | 0.2 | 0 | 0 | 0 | 0.4867 | 0.0045 | 1.1457 |
| 3 | $\cdot$ | 0.2 | 0.02 | 0 | 0 | 1.1697 | 0.0530 | 1.6987 |
| 4 | $\cdot$ | 0.2 | 0.16 | 0 | 0 | 2.0363 | 0.1738 | 2.3117 |
| 5 | $\cdot$ | 0.2 | 0.46 | 0 | 0 | 2.7937 | 0.2686 | 3.2653 |
| 6 | $\cdot$ | 0.2 | 0.69 | 0 | 0 | 3.4877 | 0.3306 | 4.1713 |
| 7 | $\cdot$ | 0.2 | 0.87 | 0 | 0 | 4.3640 | 0.4610 | 4.8990 |
| 8 | $\cdot$ | 0.2 | 0.91 | 0 | 0 | 5.5310 | 0.4584 | 5.9127 |
| 9 | $\cdot$ | 0.2 | 0.96 | 0 | 0 | 7.5303 | 0.4091 | 6.6030 |
| 10 | . | 0.2 | 1 | 0 | 0 | 10.4027 | 0.4091 | 8.2993 |

Table 2.24. Norwegian Coastal Cod. Input to the predictions using variable fishing pattern.

MFDP version 1a
Run: Pred4
Time and date: 20:37 26.04.2010
Fbar age range: 4-7

| 2010 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 21500 | 0.2 | 0 | 0 | 0 | 0.4867 | 0.0045 | 1.1457 |
| 3 | 18065 | 0.2 | 0.02 | 0 | 0 | 1.1697 | 0.0530 | 1.6987 |
| 4 | 16010 | 0.2 | 0.16 | 0 | 0 | 2.0363 | 0.1738 | 2.3117 |
| 5 | 8027 | 0.2 | 0.46 | 0 | 0 | 2.7937 | 0.2686 | 3.2653 |
| 6 | 5626 | 0.2 | 0.69 | 0 | 0 | 3.4877 | 0.3306 | 4.1713 |
| 7 | 4818 | 0.2 | 0.87 | 0 | 0 | 4.3640 | 0.4610 | 4.8990 |
| 8 | 2720 | 0.2 | 0.91 | 0 | 0 | 5.5310 | 0.4584 | 5.9127 |
| 9 | 1265 | 0.2 | 0.96 | 0 | 0 | 7.5303 | 0.4091 | 6.6030 |
| 10 | 923 | 0.2 | 1 | 0 | 0 | 10.4027 | 0.4091 | 8.2993 |
| 2011 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 21500 | 0.2 | 0 | 0 | 0 | 0.4867 | 0.0009 | 1.1457 |
| 3 | . | 0.2 | 0.02 | 0 | 0 | 1.1697 | 0.0158 | 1.6987 |
| 4 | . | 0.2 | 0.16 | 0 | 0 | 2.0363 | 0.0782 | 2.3117 |
| 5 | . | 0.2 | 0.46 | 0 | 0 | 2.7937 | 0.1868 | 3.2653 |
| 6 | . | 0.2 | 0.69 | 0 | 0 | 3.4877 | 0.3029 | 4.1713 |
| 7 | . | 0.2 | 0.87 | 0 | 0 | 4.3640 | 0.4581 | 4.8990 |
| 8 | . | 0.2 | 0.91 | 0 | 0 | 5.5310 | 0.4602 | 5.9127 |
| 9 | . | 0.2 | 0.96 | 0 | 0 | 7.5303 | 0.4102 | 6.6030 |
| 10 | . | 0.2 | 1 | 0 | 0 | 10.4027 | 0.4102 | 8.2993 |
| 2012 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 21500 | 0.2 | 0 | 0 | 0 | 0.4867 | 0.0009 | 1.1457 |
| 3 | . | 0.2 | 0.02 | 0 | 0 | 1.1697 | 0.0158 | 1.6987 |
| 4 | . | 0.2 | 0.16 | 0 | 0 | 2.0363 | 0.0782 | 2.3117 |
| 5 | . | 0.2 | 0.46 | 0 | 0 | 2.7937 | 0.1868 | 3.2653 |
| 6 | . | 0.2 | 0.69 | 0 | 0 | 3.4877 | 0.3029 | 4.1713 |
| 7 | . | 0.2 | 0.87 | 0 | 0 | 4.3640 | 0.4581 | 4.8990 |
| 8 | . | 0.2 | 0.91 | 0 | 0 | 5.5310 | 0.4602 | 5.9127 |
| 9 | . | 0.2 | 0.96 | 0 | 0 | 7.5303 | 0.4102 | 6.6030 |
| 10 | . | 0.2 | 1 | 0 | 0 | 10.4027 | 0.4102 | 8.2993 |
| 2013 |  |  |  |  |  |  |  |  |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 2 | 21500 | 0.2 | 0 | 0 | 0 | 0.4867 | 0.0009 | 1.1457 |
| 3 | . | 0.2 | 0.02 | 0 | 0 | 1.1697 | 0.0158 | 1.6987 |
| 4 | . | 0.2 | 0.16 | 0 | 0 | 2.0363 | 0.0782 | 2.3117 |
| 5 | . | 0.2 | 0.46 | 0 | 0 | 2.7937 | 0.1868 | 3.2653 |
| 6 | . | 0.2 | 0.69 | 0 | 0 | 3.4877 | 0.3029 | 4.1713 |
| 7 | . | 0.2 | 0.87 | 0 | 0 | 4.3640 | 0.4581 | 4.8990 |
| 8 | . | 0.2 | 0.91 | 0 | 0 | 5.5310 | 0.4602 | 5.9127 |
| 9 | . | 0.2 | 0.96 | 0 | 0 | 7.5303 | 0.4102 | 6.6030 |
| 10 | . | 0.2 | 1 | 0 | 0 | 10.4027 | 0.4102 | 8.2993 |

Table 2.25. Norwegian Coastal Cod. Predictions using fixed fishing pattern.

MFDP version 1a
Run: Pred1
Norwegian Coastal Cod
Time and date: 20:10 26.04.2010
Fbar age range: 4-7

2010

| Biomass | SSB | FMult | FBar | Landings |
| :--- | :--- | :--- | :--- | :--- |
| 161439 | 80222 | 1 | 0.3085 | 36815 |


| 2011 |  |  |  | 2012 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 159005 | 78871 | 0.4 | 0.1234 | 15621 | 177543 | 94316 |
| . | 78871 | 0.55 | 0.1697 | 20973 | 171856 | 89803 |
| . | 78871 | 0.7 | 0.216 | 26074 | 166451 | 85533 |
| . | 78871 | 0.85 | 0.2622 | 30935 | 161311 | 81492 |
| . | 78871 | 1 | 0.3085 | 35570 | 156422 | 77667 |
| . | 78871 | 1.15 | 0.3548 | 39991 | 151771 | 74047 |
| . | 78871 | 1.3 | 0.4011 | 44209 | 147344 | 70618 |

MFDP version 1a
Run: Pred2
Pred2MFDP Index file 26.04.2010
Time and date: 20:23 26.04.2010
Fbar age range: 4-7

2010

| Biomass | SSB | FMult | FBar | Landings |
| :--- | :--- | :--- | :--- | :--- |
| 161439 | 80222 | 1 | 0.3085 | 36815 |

2011

| Biomass | SSB | FMult | FBar | Landings |
| :--- | :--- | :--- | :--- | :--- |
| 159005 | 78871 | 0.85 | 0.2622 | 30935 |


| 2012 |  |  |  | 2013 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 161311 | 81492 | 0.4 | 0.1234 | 15839 | 178711 | 96035 |
| . | 81492 | 0.55 | 0.1697 | 21270 | 172980 | 91463 |
| . | 81492 | 0.7 | 0.216 | 26447 | 167531 | 87135 |
| . | 81492 | 0.85 | 0.2622 | 31383 | 162347 | 83036 |
| . | 81492 | 1 | 0.3085 | 36091 | 157415 | 79154 |
| . | 81492 | 1.15 | 0.3548 | 40582 | 152720 | 75477 |
| . | 81492 | 1.3 | 0.4011 | 44869 | 148250 | 71992 |

Input units are thousands and kg - output in tonnes

Table 2.26. Norwegian Coastal Cod. Predictions using variable fishing pattern.

MFDP version 1a
Run: Pred3
Pred1MFDP Index file 26.04.2010
Time and date: 20:34 26.04.2010
Fbar age range: 4-7

2010

| Biomass | SSB | FMult | FBar | Landings |
| :--- | :--- | :--- | :--- | :--- |
| 161439 | 80222 | 1 | 0.3085 | 36815 |


| 2011 |  |  |  | 2012 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 159005 | 78871 | 0.4 | 0.1026 | 12958 | 180280 | 95719 |
| . | 78871 | 0.55 | 0.1411 | 17397 | 175533 | 91678 |
| . | 78871 | 0.7 | 0.1796 | 21627 | 171023 | 87854 |
| . | 78871 | 0.85 | 0.218 | 25658 | 166735 | 84233 |
| . | 78871 | 1 | 0.2565 | 29501 | 162658 | 80804 |
| . | 78871 | 1.15 | 0.295 | 33168 | 158780 | 77555 |
| . | 78871 | 1.3 | 0.3335 | 36668 | 155090 | 74476 |

MFDP version 1a
Run: Pred4
Pred3MFDP Index file 26.04.2010
Time and date: 20:37 26.04.2010
Fbar age range: 4-7

2010

| Biomass | SSB | FMult | FBar | Landings |
| :--- | :--- | :--- | :--- | :--- |
| 161439 | 80222 | 1 | 0.3085 | 36815 |
|  |  |  |  |  |
| 2011 |  |  |  |  |
| Biomass | SSB | FMult | FBar | Landings |
| 159005 | 78871 | 1 | 0.2565 | 29501 |


| 2012 |  |  |  | 2013 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 162658 | 80804 | 0.4 | 0.1026 | 13183 | 183053 | 97853 |
| . | 80804 | 0.55 | 0.1411 | 17711 | 178276 | 93804 |
| . | 80804 | 0.7 | 0.1796 | 22030 | 173731 | 89965 |
| . | 80804 | 0.85 | 0.218 | 26152 | 169405 | 86325 |
| . | 80804 | 1 | 0.2565 | 30088 | 165285 | 82872 |
| . | 80804 | 1.15 | 0.295 | 33846 | 161361 | 79596 |
| . | 80804 | 1.3 | 0.3335 | 37438 | 157622 | 76486 |

Input units are thousands and kg - output in tonnes


Figure 2.1. Norwegian statistical rectangles in the Barents Sea. Coastal cod catches are estimated from the total cod catch taken inside 12 n.mile in areas 03 and 04 . The same areas are also referred to in the survey results (sec. 2.3).


Figure 2.2. Norwegian statistical rectangles in the Norwegian Sea. Coastal cod catches are estimated from the total cod catch taken inside 12 n.mile in areas $05,00,06$ and 07 . The same areas are also referred to in the survey results (sec. 2.3).


Figure 2.3. Map showing Vestfjorden, the Norwegian statistical area 00 ("OMRADE 00") with the south-western location 03 and 04 and the north-eastern locations 46 and 48.


Figure 2.4. Estimated landings of Norwegian coastal cod. Commercial landings in blue and recreational catches in red.


Figure 2.5. An image of a coastal cod otolith (top) and a north-east Arctic cod otolith (bottom). The two first translucent zones are highlighted. (from Berg et al. 2005)


Figure 2.6 Coastal cod. Abundance at age in the total survey.
Upper: ages 2-5, Lower: ages 6-8.


Figure 2.7 Coastal cod. Abundance at age in the survey, statistical area 03.

Upper: ages 2-5, Lower: ages 6-8.



Figure 2.8 Coastal cod. Abundance at age in the survey, statistical area 04.

Upper: ages 2-5, Lower: ages 6-8.


Figure 2.9 Coastal cod. Abundance at age in the survey, statistical area 05.

Upper: ages 2-5, Lower: ages 6-8.


Figure 2.10 Coastal cod. Abundance at age in the survey, statistical rectangle 00.
Upper: ages 2-5, Lower: ages 6-8.


Figure 2.11 Coastal cod. Abundance at age in the survey, statistical area 06.
Upper: ages 2-5, Lower: ages 6-8



Figure 2.12 Coastal cod. Abundance at age in the survey, statistical area 07.

Upper: ages 2-5, Lower: ages 6-8.


Figure 2.13 Coastal cod. Different biomass estimates from the Norwegian coastal survey: Total biomass, 4+ biomass, and spawning biomass (left axis). Also the SSB from the final VPA is included (right axis).


Figure 2.14. Relative harvest rate; Yield relative to the 4+ biomass estimated from the survey (left axis). Also added the F4-7 from the final VPA (right axis).


Figure 2.15 Norwegian coastal cod (NCC). Retrospective plots for assessment years 2000-2010 using standard settings except stock dependent catchability for ages 2 and 3 , and keeping weight, maturity and natural mortality as estimated in 2010 for all runs.

## 3 North-East Arctic Cod (Subareas I and II)

### 3.1 Status of the fisheries

### 3.1.1 Historical development of the fisheries (Table 3.1a)

From a level of about 900,000 $t$ in the mid-1970s, landings declined steadily to around $300,000 \mathrm{t}$ in 1983-1985 (Table 3.1a). Landings increased to above $500,000 \mathrm{t}$ in 1987 before dropping to $212,000 \mathrm{t}$ in 1990, the lowest level recorded in the post-war period. The catches increased rapidly from 1991 onwards, stabilized around 750,000 t in 19941997 but decreased to about $414,000 \mathrm{t}$ in 2000. After 2000, the reported catches have been between 400,000 and $520,000 \mathrm{t}$, in addition there have been unreported catches (see below). The fishery is conducted both with an international trawler fleet and with coastal vessels using traditional fishing gears. Quotas were introduced in 1978 for the trawler fleets and in 1989 for the coastal fleets. In addition to quotas, the fishery is regulated by a minimum catch size, a minimum mesh size in trawls and Danish seines, a maximum by-catch of undersized fish, closure of areas having high densities of juveniles and by seasonal and area restrictions.

### 3.1.2 Reported landings prior to 2010 (Tables 3.1-3.3, Figure 3.1)

Reported landings of cod in subarea I and Divisions IIa and IIb:
Final official landings for 2008 amount to $462,364 \mathrm{t}$. The provisional landings for 2009 reported to the w orking group are $538,660 \mathrm{t}$.

Reported landings figures used for the assessment of North-East Arctic cod:
The historical practice (considering catches between $62^{\circ} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ for the whole year and catches between $67^{\circ} \mathrm{N}$ and $69^{\circ} \mathrm{N}$ for the second half of the year to be Norwegian coastal cod) leads to reported landings of North-East Arctic cod of 449,171 t in 2008 and 523,431 t in 2009 (Table 3.3). The coastal cod catches calculated this way in 2008 and 2009 were 13,193 t and 15,229 t, respectively. The catches of coastal cod calculated this way for the period 1960-2009 are given in Table 3.1b together with the coastal cod catches calculated based on otolith types as described in Section 2.
The landings by area, are shown in Table 3.1a, and further split into trawl and other gears in Table 3.2. The distribution of catches by areas and gears in 2009 was similar to 2008. The nominal landings by country are given in Table3.3.

There is information on cod discards (see section 0.5) but it was not included in the assessment because this data are fragmented and different estimates are in contradiction with each other. Moreover the level of discards is relatively small in recent period and inclusion of these estimates in the assessment should not change our perception on NEA cod stock size.

### 3.1.3 Unreported catches of Northeast Arctic cod in 2002-2009

In the years 2002-2008 certain quantities of unreported catches (IUU catches) have been added to the reported landings. More details on this issue are given in Section 0.4. The Norw egian and Russian estimates of IUU for this period are given in Table 3.1a. For 2009 there was a report from the Norwegian-Russian analytical group on estimation of total catches. According to that report the total catches of cod are very close (within 1\%) to officially reported landings. The Working Group decided not to include IUU catches in 2009.

### 3.1.4 TACs and advised catches for 2009 and 2010

The Joint Norwegian-Russian Fisheries Commission (JNRFC) agreed on a cod TAC of $546,000 \mathrm{t}$ for 2009, including $21,000 \mathrm{t}$ Norw egian coastal cod. The total reported catch of $538,660 \mathrm{t}$ in 2009 was $7,340 \mathrm{t}$ below the agreed TAC.
The advice for 2010 given by ACFM in 2009 w as based on the assessment made by AFWG in 2009. The agreed harvest control rule then implied a NEA cod TAC for 2010 of 577,500 tonnes. However, the JNRFC made an amendment to the agreed rule (see section 3.6.3), introducing a lower bound on $\mathrm{F}(0.30)$ when the spawning stock biomass is above $\mathrm{B}_{\mathrm{pa}}$. This amended rule gave a NEA cod TAC for 2010 of 607,000 tonnes, which was the quota set by JNRFC for 2010. In addition, the TAC for Norwegian Coastal Cod was set to the same value for 2010 as for 2009: 21,000 t.
The Working Group has no information on the size of expected unreported landings in 2010.

### 3.2 Status of research

### 3.2.1 Fishing effort and CPUE (Table A1)

Updated CPUE series of the Norwegian and Russian trawl fisheries are given in Table A1. The data reflect the total trawl effort, both for Norway and Russia. The Norwegian series is given as a total for all areas (Table A1).

### 3.2.2 Survey results - abundance and size at age (Tables 3.6, A2-A14)

Joint Barents Sea winter survey (bottom trawl and acoustics)
The preliminary swept area estimates and acoustic estimates from the Joint winter survey on demersal fish in the Barents Sea in winter 2010 are given in Tables A2 and A3. More details on this survey are given in Aglen et al. (WD 15). The coverage was fairly good within the strata system defined for the survey. There has been a pattern in recent years to have concentrations of cod near the borders of the strata system. This could indicate an increasing amount of fish being distributed outside the strata system.

Before 2000 this survey was made without participation from Russian vessels, while in 2001-2005 and 2008-2010 Russian vessels have covered important parts of the Russian zone. In 2006-2007 the survey was carried out only by Norwegian vessels. In 2007 the vessels were not allowed to cover the Russian EEZ. The method for adjustment for incomplete area coverage in 2007 is described in the 2007 report. Table 3.6 shows areas covered in the time series and the additional areas implied in the method used to adjust for missing coverage in Russian Economic Zone. In 4 of the 5 adjusted years the adjustments were not based on area ratios, but the "index ratio by age" was used. This means that the index by age (for the area outside REZ) was scaled by the observed ratio between total index and the index outside REZ observed in the years prior to the survey.
Regarding the older part of this time series it should be noted that the survey prior to 1993 covered a smaller area (Jakobsen et al.1997), and the number of young cod (particularly 1- and 2-year old fish) was probably underestimated. Other changes in the survey methodology through the time are described by Jakobsen et al. (1997). Note that the change from 35 to 22 mm mesh size in the codend in 1994 is not corrected for in the time series. This mainly affects the age 1 indices.

## Lofoten acoustic survey on spawners

The estimated abundance indices from the Norwegian acoustic survey off Lofoten and Vesterålen (the main spawning area for this stock) in March/April are given in Table A4. A description of the survey, sampling effort and details of the estimation procedure can be found in Korsbrekke (1997). The 2010 survey showed about 20\% increase in numbers compared to the 2009 survey, while the biomass was similar. The percentage of repeat spawners was $50 \%$, compared to $49 \%$ in 2009.

## Russian autumn survey

Abundance estimates from the Russian autumn survey (November-December) are given in Table A9 (acoustic estimates) and Table A10 (bottom trawl estimates). The entire bottom trawl time series was in 2007 revised backwards to 1982 (Golovanov et al., 2007, WD3), using the same method as in the revision presented in 2006, which went back to 1994. The new swept area indices reflect Northeast Arctic cod stock dynamics more precisely compared to the previous one - catch per hour trawling. The Russian autumn survey in 2006 was carried out with reduced area coverage. Divisions IIa and Ilb were adequately investigated in the survey in contrast to Sub-area I, where the survey covered approximately $40 \%$ of the long-term average area coverage. The Subarea I survey indices were calculated based on actual covered area (40 541 sq. miles). The 2007 AFWG decided to use the final year class indices without any correction because of satisfactory internal correspondence between year class abundances at age 2-9 years according to the 2006 survey and ones due to the previous surveys.

The Russian autumn 2009 survey was conducted in the standard period and under the standard methods. An area of $203{ }^{*} 10^{3}$ sq. miles was covered, which is somewhat larger than the standard area. The 2009 abundance indices were calculated based on the standard area adopted at the two previous AFWG (2007 and 2006) (Golovanov et al.,WD3 in 2007; WD 21 in 2006).

Overall increase of cod numbers was observed in the last survey, especially for cod at age 5-7 and for ages 9-10. Estimates for ages 9 and 10 were the highest ones over the time series. Rather wide distribution of cod was registered, and besides, delaying of return migrations of maturing fish from the eastern feeding grounds was observed.

## Joint Ecosystem survey

Swept area bottom trawl estimates from the joint Norwegian-Russian ecosystem survey in August-September for the period 2004-2009 are given in Table A14. The new index values were calculated in 2010 (WD 20). This time series have been tested as new tuning fleet in XSA (section 3.11.3). Survey results - length and weight at age (Tables A5-A8, A11-A12)

Length at age is shown in Table A5 for the Norwegian survey in the Barents Sea in winter, in Table A7 for the Lofoten survey and in Table A11 for the Russian survey in October-December. Weight at age is shown in Table A6 for the Norwegian survey in the Barents Sea in winter, in Table A8 for the Lofoten survey and in Table A12 for the Russian survey in October-December.
Both the Joint winter survey in 2010 and the Russian autumn survey in 2009 show a continued slight tendency on reduction of size-at-age compared to the previous surveys (Table A6 and A12).

### 3.2.3 Age reading

The joint Norwegian-Russian work on cod otolith reading has continued, with regular exchanges of otoliths and age readers (see chapter 0.5). The results of fifteen years of annual comparative age readings are described in Yaragina et al. (2009b). Zuykova et al. (2009) re-read old otoliths and found no significant difference in contemporary and historical age determination and subsequent length at age. However, age at first maturation in the historical material as determined by contemporary readers is younger than that determined by historical readers. Taking this difference into account would thus have effect on the spawning stock-recruitment relationship and thus on the biological reference points.

### 3.3 Data used in the assessment

### 3.3.1 Catch at age (Tables 3.7 and 3.9)

For 2009, age compositions from all areas were available from Russia, Germany and Norway. Spain provided age compositions from Divisions IIa and Ilb. Poland provided age compositions from Division Ilb. Unsampled catches were distributed on age by using data from Russian trawl in Sub-area I and Division IIa, and by using data from Norwegian trawl in Division IIb. Table 3.7 shows available catch at age data for all ages 1-15+. The 2009 catch at age data was calculated using Intercatch, see section 0.9 .

### 3.3.2 Weight at age (Tables 3.4 and 3.10-3.11).

## Catch weights

For 2009, the mean weight at age in the catch (Table 3.10) was obtained from Intercatch as a weighted average of the weight at age in the catch for Norway, Russia, Germany, Spain and Poland. The weight at age in the catch for these countries is given in Table 3.4.

## Stock weights

Since ages 12 and 13+ are scarce in the survey samples, fixed values for these ages have formerly been used (set equal to typical weights for these ages observed in catches). Since the 2000 working group the assessment has applied 13 as plus group. For the years 1946-1984 the 13+ weights are calculated year by year as a weighted mean of the former fixed values for older ages. For later years they are calculated from the average observed weight for age 11 in the years 1995-2008 increased by 1.58 kg for age 12 and $2 \times 1.58 \mathrm{~kg}$ for age $13+$.

For ages 1-11 stock weights at age at the start of year y ( $W_{a, y}$ ) for 1983-2010 (Table 3.12) were calculated as follows:
$W_{a, y}=0.5\left(W_{r u s, a-1, y-1}+\left(\frac{N_{\text {nbar }, a}, y W_{\text {nbar }, a}, y+N_{\text {lof }, a, y}, W_{\text {lof }, a, y}}{N_{\text {nbar }, a}, y+N_{\text {lof }, a}, y}\right)\right)$
where
$W_{\text {rusa-1,y-1 }}$ : Weight at age a-1 in the Russian survey in year y-1 (Table A12)
$N_{n b a r a, y}$ : Abundance at age a in the Norwegian Barents Sea acoustic survey in year y (Table A2)
$W_{\text {rbara, } y}$ : Weight at age a in the Norwegian Barents Sea acoustic survey in year y (Table A6)
$N_{l o f a, y}$ : Abundance at age a in the Lofoten survey in year y (Table A4)
$W_{b f a, y}:$ Weight at age a in the Lofoten survey in year y (Table A8)

### 3.3.3 Natural mortality

A natural mortality of 0.2 was used. In addition, cannibalism was taken into account as described in Section 3.4.2. The proportion of F and M before spawning was set to zero.

### 3.3.4 Maturity at age (Tables 3.5 and 3.12 )

Historical (pre 1982) Norwegian and Russian time series on maturity ogives were reconstructed by the 2001 AFWG meeting (ICES CM 2001/ACFM:19). The Norwegian maturity ogives were constructed using the Gulland method for individual cohorts, based on information on age at first spawning from otoliths. For the time period 1946-1958 only the Norwegian data were available. The Russian proportions mature at age, based on visual examinations of gonads, were available from 1959.

Since 1982 Russian and Norwegian survey data have been used (Table 3.5). For the years 1985-2010, Norwegian maturity at age ogives have been obtained by combining the Barents Sea winter survey and the Lofoten survey. Russian maturity ogives from the autumn survey as well as from commercial fishery for November-February are available from 1984 until present. The Norwegian maturity ogives tend to give a higher percent mature at age compared to the Russian ogives, which is consistent with the generally higher growth rates observed in cod sampled by the Norwegian surveys. The approach used is consistent with the approach used to estimate the weight at age in the stock (described in Section 3.3.2). The percent mature at age for the Russian and Norwegian surveys have been arithmetically averaged for all years, except 1982-1983 when only Norw egian observations were used and 1984 when only Russian observations were used.

### 3.3.5 Cannibalism

The method used for calculation of the prey consumption by cod described by Bogstad and Mehl (1997) is used to calculate the consumption of cod by cod for use in XSA. The consumption is calculated based on cod stomach content data taken from the joint PINRO-IMR stomach content database (methods described in Mehl and Yaragina 1992). On average about 9,000 cod stomachs from the Barents Sea have been analyzed annually in the period 1984-2009. The consumption calculations this year have been updated by data for 2009. Also, the data for 2004-2008 have been revised, as it was discovered that data from the southeastern corner of the Barents Sea (east of $50^{\circ} \mathrm{E}$ and south of $74^{\circ} \mathrm{N}$ ) were not included in previous calculations. These data are used to calculate the per capita consumption of cod by cod for each half-year (by prey age groups $0-6$ and predator age groups 1-11+). It was assumed that the mature part of the cod stock is found outside the Barents Sea for three months during the first half of the year. Thus, consumption by cod in the spawning period was omitted from the calculations.

The number of cod predators at age is taken from the VPA, and thus an iterative procedure has to be applied (Section 3.4.2). All occurrences of intra-cohort predation
were removed from the data set as these could possibly cause problems with convergence.

### 3.4 Assessment using VPA models

The XSA was also this year used as the main assessment method, as an update assessment was carried out. The TISVPA method was also run on the same data. Additional assessment methods (survey calibration of VPA and Gadget) are presented in Section 3.9.

The following surveys and commercial CPUE data series were used for tuning of both models:

| XSA <br> name | TISVPA <br> name | Name | Place | Season | Age | Years |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Fleet 09 | Fleet1 | Russian trawl <br> CPUE | Total area | Allyear | $9-11$ | 1985-2009 |
| Fleet 15 | Fleet2 | Joint bottom <br> trawl survey | Barents Sea | Feb-Mar | $3-8$ | $1981-2010$ |
| Fleet 16 | Fleet3 | Joint acoustic <br> survey | Barents <br> Sea+Lofoten | Feb-Mar | $3-9$ | $1985-2010$ |
| Fleet 18 | Fleet4 | Russian <br> bottom trawl <br> surv. | Total area | Oct-Dec | $3-9$ | 1994-2009 |

As in earlier assessments the surveys that were conducted during winter were allocated to the end of the previous year. This was done so that data from the surveys in 2010 could be included in the assessment. The tuning fleet file is shown in Table 3.13. Note that the joint acoustic survey (sum of Barents Sea and Lofoten acoustic survey indices) is given in Table A13.

### 3.4.1 XSA settings

The output tables from the tuning include ages 1 and 2, just to show the year class abundance at age 1 and 2 created by the cannibalism numbers (Section 3.4.3). These age groups are not included in the tuning, however.

Some of the survey indices have been multiplied by a factor 10. This was done to keep the dynamics of the surveys even for very low indices, because XSA adds 1.0 to the indices before the logarithm is taken.

XSA was run using default settings with the following exceptions:
Tapered time weighting power 3 over 10 years
Catchability dependent of stock size for ages less than 6
F of the final 5 years and the 2 oldest age groups used in F shrinkage
Standard error of the mean to which estimates are shrunk set to 1.0
These settings are identical to those used by last years' Working Group. Since the assessments in August 2000, few changes in model settings and data choices have been made.

### 3.4.2 Including cannibalism in XSA (Table 3.8)

The catch numbers shown in Table 3.9 together with cannibalism numbers (Tables 3.8) were used in the XSA tuning.

For the cod assessment data from annual sampling of cod stomachs has been used for estimating cannibalism, since the 1995 assessment. The argument has been raised that the uncertainty in such calculations are so large that they introduce too much noise in the assessment. A rather comprehensive analysis of the usefulness of this was presented in Appendix 1 in the 2004 AFWG report. The conclusion was that it improves the assessment.

The following procedure was followed: As a starting point the number of cod consumed by cod was estimated from the stock estimates in the last assessment and the per capita estimates of consumption of cod by cod. Then the number consumed was added to the catches used for tuning. The resulting stock then leads to new estimates of consumption. This procedure was repeated until the consumed numbers for the latest year (2009) differed less than $1 \%$ from the previous iteration. The final numbers of cod eaten by cod are given in Table3.8.

It would be promising to include cannibalism to the historical period (1946-1983) data to make the VPA time series consistent. There have been some approaches proposed (Yaragina et al. 2009a).

### 3.4.3 XSA tuning diagnostics (Table 3.14-3.15, Figure 3.2-3.3)

Thetuning diagnostics from XSA with cannibalism are given in Table3.15. Figure 3.2 shows the log catchability residuals of the tuning series. It is observed a slight positive trend in residuals of the winter bottom trawl survey (Fleet 15) for ages 6-8. Most of the residuals are negative in 2006 and positive in 2007 for the combined winter + Lofoten acoustic survey (Fleet 16). The residuals in 2009 are close to zero and have no particular pattern. For age 9 and 10 in fleet 09 (Russian commercial CPUE) there seem to be big negative (in 2008) and positive (in 2009) residuals.

Figure 3.3 and Table 3.14 compares the estimated survivors (by end of 2009) and Fs before shrinkage in single fleet tunings. (The single fleet runs applies the same shrinkage settings as the standard run, but the tabulated values of F and survivors are the pure survey predictions in the diagnostics output). Survivors' estimates from single fleet runs for all ages are in a fair agreement between fleets. Nevertheless, final XSA run including all fleets tends to give higher estimates of survivors at ages 3, 5-7 compare to single fleet runs. This could be explained by higher influence of shrinkage in single fleet runs.

ACFM technical minutes have several times commented on the rather unconventional use of "stock size dependant catchability" (ssdq). For NEA cod, this is assumed for age groups 3-5. It is true that this choice involves more parameters to be estimated and a likely less precise parameter fit, in particular when the tuning is restricted to the latest 10 years. It is also observed that the influence of shrinkage is considerably higher for the age groups estimated by this $q$-assumption (Table 3.14). The 2005 WG argued for keeping this setting on the basis of compared retrospective patterns, and the ACFM reviewers agreed that without ssdq some problems might occur again as soon as some high survey values occur. In spite of rather high survey values for ages 3 and 4 in the previous two years, a test run without stock size dependant catchability gave slightly lower F and higher SSB for 2009.

Several earlier assessments have shown to be sensitive both to the length of the tuning period, and the choice of stock size dependant catchability. The following comparative runs were made to explore the sensitivity to these choices:

| Model setting different from the standard run | $F(5-10)-09$ | SSB-09 |
| :--- | :--- | :--- |
| No stock size dependant catchability | 0.253 | 1153 |
| Stock size dependant catchability for ages 3-7 | 0.301 | 987 |
| 15year tuning period | 0.260 | 1123 |
| Minimum SE for fleet weighting reduced from 0.3 to 0.1 | 0.276 | 1089 |
| Standard run | 0.276 | 1077 |

From this it seems that the assessment is rather robust to the above changes of model setting, but the diagnostics discussed above indicate some sensitivity to the choice of tuning data.
Retrospective plots of F, SSB and recruitment, going back to 2000 as the last year in the assessment, are shown in Figure 3.4. Cannibalism is taken into account, but the number of cod consumed by cod was not recalculated year by year in the retrospective analysis. The retrospective pattern seems satisfactory.

### 3.4.4 Results (Table 3.16-3.26)

The total fishing mortalities (true fishing mortality plus mortality from cannibalism) and population numbers are given in Tables 3.16 and 3.17.
In order to build a matrix of natural mortality which includes predation, the fishing mortality estimated in the final XSA analyses was split into the mortality caused by the fishing fleet (real F ) and the mortality caused by cod cannibalism (M2 in MSVPA terminology) by using the number caught by fishing and by cannibalism. The new natural mortality matrix was prepared by adding 0.2 (M1) to the M2. This new M matrix (Table 3.18) was used together with the new real Fs (Table 3.20) to run the final VPA on ages 3-13+. M2 and F values for ages 1-6 in 1984-2009 are given in Tables 3.19 and 3.21.

The stock numbers from the final run are given in Tables 3.22, while the corresponding stock biomass at age and the spawning stock biomass at age are given in Tables 3.23-3.24. Summaries of landings, fishing mortality, stock biomass, spawning stock biomass and recruitment since 1946 runs are given in Table 3.25 and Figure 3.1.

Cannibalism on cod age 3 and older may of course also have occurred before 1984. Thus, there is an inconsistency in the recruitment time series. For comparison with the historic time series an additional VPA with the same terminal Fs and fixed natural mortality (0.2) is presented (Table 3.26).

### 3.4.5 TISVPA (Fig 3.5-3.8)

The TISVPA (Triple Instantaneous Separable VPA) model (Vasilyev, 2006) represents fishing mortality coefficients (more precisely - exploitation rates) as a product of three parameters: $\mathrm{f}(\mathrm{year})^{*} \mathrm{~s}(\mathrm{age})^{*} \mathrm{~g}($ cohort $)$. The generation-dependent parameters, which are estimated within the model, are intended to adapt traditional separable representation of fishing mortality to situations when several year classes may have peculiarities in their interaction with fishing fleets caused by different spatial distribution, higher attractiveness of more abundant schools to fishermen, or for some other reasons. The model was first presented and tested at the ICES Working Group on Methods of Fish Stock Assessments (WGMG 2006) and was used for stock assessment
for several ICES stocks, including North-East Atlantic mackerel, blue whiting, Norwegian spring spawning herring (WGMHSA 2006, 2007; WGNPBW 2006, 2007) and for NEA cod (AFWG 2008 and 2009). The model is an extension of the ISVPA model (Kizner and Vasilyev, 1997; Vasilyev, 2005).
This year the TISVPA model was applied to NEA cod data too. Natural mortality from cannibalism, variable by age and year values were taken from the XSA runs as. As well as in XSA runs, in trial runs 4 sets of age-structured tuning data were included into analysis: Russian trawl cpue ("fleet 1"); joint bottom trawl surveys ("fleet 2 "); joint acoustic surveys (Barents Sea and Lofoten) - "fleet 3", and Russian bottom trawl surveys ("fleet 4").
Settings of the TISVPA model were similar to the previous year assessment, but with some corrections made in order to make more clear the signals from different sources of information: so called "mixed" version, reserving the possibility of errors both in catch-at-age and in separable representation. Additional restriction on the solution was unbiased model description of logarithmic catch-at-age data.

TISVPA model version allows us to include or not the cohort factor into calculations. The generation-dependent factors in triple-separable representation of fishing mortality coefficients were estimated for age groups 4 to 11 to exclude most noisy age groups. The trial run showed that this factor differs significantly from 1, which is the reason for including it in the final run.

The experiments with various versions of loss function component for catch-at age matrix showed that AMD (absolute median deviations) had to be chosen as in the this case minimum was more pronounced.

Preliminary experiments revealed year-specific patterns in residuals for fleets 1 and 4 , perhaps, due to different surveys conditions. That is why for these fleets not abun-dances-at-age, but age proportions were tuned.

For all "fleets", except fleet 4, for which AMD gave a more clear minimum, the simplest measure of closeness of fit of the model to the data - sum of squares of residuals in logarithmic abundance-at-age gave the apparent minima of respective components of the model loss function (Figure3.5).

Residuals in logarithmic catch-at-age and in abundance-at-age (for fleets) are shown on Fig.3.6.

The retrospective analysis was carried out with the same options as the final TISVPA run. Results of retrospective runs (Figure 3.7) show a reasonable historical stability of the estimates and the absence of systematic shifting tendency.

It is necessary to underline that extremely high estimates of abundance at age 3 in 2007 and 2008 in the results are due to high catch of these age groups as the abundance estimates are directly come from the catch value and average selection. These estimates are always the least reliable ones. Figure 3.6 for $C(a, y)$ demonstrates that residuals in catch-at-age data for age 3 are higher than for other ages. Figure 3.8 compares the estimates of abundance at age 3 for previous years obtained with help of TISVPA to historical catches at age 3 (upper figure) and to index abundance- Fleet 2 (lower figure). As it can be seen, the TISVPA recruitment values at age 3 in 2007-2008 were only a little higher (and in 2009 they are lower), than in 1986.
There are a number of properties of the TISVPA model which make the model a valuable tool for data exploration in NEA cod stock assessment. These properties include the possibility to strictly formulate a statistical meaning of the solution; not to
consider as absolutely true the catch-at-age data, survey data, fleet cpue, or the assumption about stability of selection pattern; to take into account the generationdependent peculiarities in selection pattern; to trace the information about the stock size independently from each source of data (including catch-at-age); attention to robustness of the results (by means of possibility to apply robust measures of the goodness of fit and to ensure the unbiasedness of the solution), as well as the experience in its application to other ICES stocks.
The total stock biomass in 2009 from the TISVPA runs totaled 3,2 million tons, while the spawning stock biomass was 1,145 million tons and $\mathrm{F} 5-10$ in 2009 was 0.26 .

### 3.4.6 Comparis on of TISVPA and XSA results (Fig 3.9)

A comparison of the results from the TISVPA and XSA are given in Figure 3.9. The trends are similar as seen from the plots. The difference is remarkable after 2006, in terminal (2009) year TISVPA estimates of total stock biomass are higher by about $20 \%$, spawning stock biomass - by $6 \%$ as compare with XSA.
The TISVPA run gives an $F(5-10)$ for 2009 of 0.26 , while the XSA gives 0.28 .

### 3.5 Results of the ass essment

### 3.5.1 Fishing mortalities and VPA (Tables 3.20-3.25, Figure 3.1)

The estimated $\mathrm{F}_{5-10}$ in 2009 from the SVPA is 0.28 , which is below $\mathrm{F}_{\mathrm{pa}}$ and is the lowest since 1990 . Fishing mortality has gradually declined since 2005 . The spawning stock biomass in 2010 is estimated to be $1,145,000 \mathrm{t}$, which is the highest since 1947. Total stock biomass in 2010 is estimated to $2,645,000$ tonnes which is not that outstanding in the time series. One should bear in mind that in the early part of the time series the fraction mature was lower.

### 3.5.2 Recruitment (Table 1.13)

Since survey data for the youngest ages are not used in the XSA, these ages are estimated by other models. At the 2008 it was decided to use a hybrid model, which is an arithmetic mean of different recruitment models (Section 1.4.5). It was agreed to use the same approach this year. The input data for those models are the following time series; survey data for ages 0,1 and 2 (Russian autumn survey) and ages 1,2 and 3 (Joint winter survey), 0 -group from the ecosystem survey, capelin biomass, ice coverage, temperature and oxygen saturation at the Kola section, air temperature at Murman coast. Prognosis from all the models, including the hybrid is presented in Table 1.13. Here also the results from the earlier used RCT3 model are shown. The numbers at age 3 calculated by the hybrid method were: 384 million for the 2007 year class, 465 million for the 2008 year class and 484 million for the 2009 year class.

### 3.6 Reference points and harvest control rules

New reference points for Northeast Arctic cod were proposed by SGBRP in January 2003 (ICES CM 2003/ACFM:11) and adopted by ACFM at the May 2003 meeting.
At the 38 session of JRNFC the NEA cod HCR has been revised and a new version of the management rule was adopted (see section 3.6.3). The new HCR has been evaluated during the current meeting and considered to be in accordance with precautionary approach.

In according to the request from the Norwegian Ministry of Fishery and Coastal Affairs, TAC advice for 2011 is based on the new rule.

### 3.6.1 Biomass reference points (Figure 3.1)

The values adopted by ACFM in 2003 are $\mathbf{B}_{\lim }=220,000 \mathrm{t}, \mathbf{B}_{\mathrm{pa}}=460,000 \mathrm{t}$. (ICES CM 2003/ACFM:11).

### 3.6.2 Fishing mortality reference points

The values adopted by ACFM in 2003 are $\mathbf{F}_{\text {lim }}=0.74$ and $\mathbf{F}_{\mathrm{pa}}=0.40$. (ICES CM 2003/ACFM:11).

Calculations of yield per recruit gave the following values: $\mathrm{F}_{0.1}=0.15$ and $\mathrm{F}_{\max }=0.28$.

### 3.6.3 Harvest control rule

At the 31 st session of The Joint Norwegian-Russian Fishery Commission (JRNFC) in autumn 2002, the Parties agreed on a new harvest control rule. This rule was applied for the first time when setting quotas for 2004. The rule was somewhat amended at the 33 rd session of The Joint Norwegian-Russian Fishery Commission in autumn 2004. The amended rule was evaluated by ICES in 2005 and found to be precautionary.
"The Parties agreed that the management strategies for cod and haddock should take into account the following:
conditions for high long-term yield from the stocks
achievement of year-to-year stability in TACs
full utilization of all available information on stock development
On this basis, the Parties determined the following decision rules for setting the annual fishing quota (TAC) for Northeast Arctic cod (NEA cod):
estimate the average TAC level for the coming 3 years based on $F_{p a}$. TAC for the next year will be set to this level as a starting value for the 3-year period.
the year after, the TAC calculation for the next 3 years is repeated based on the updated information about the stock development, however the TAC should not be changed by more than $+/-10 \%$ compared with the previous year's TAC.
if the spawning stock falls below Bpa, the procedure for establishing TAC should be based on a fishing mortality that is linearly reduced from $F_{p a}$ at $B_{p a}$, to $F=0$ at SSB equal to zero. At SSB-levels below Bpa in any of the operational years (current year, a year before and 3 years of prediction) there should be no limitations on the year-to-year variations in TAC.

A review and discussion of this and other harvest control rule was made by the ICES SGMAS (ICES 2007c). They discovered that this HCR may give unexpected and possibly unwanted results if the assessment changes much from year to year in a situation when SSB is close to $\mathrm{B}_{\text {pa. }}$. This problem has, however, so far not been encountered in the application of the HCR.
At the $38^{\text {th }}$ JNRFC meeting, an amendment was made to the rule, and it now reads (new text in bold):
"On this basis, the Parties determined the following decision rules for setting the annual fishing quota (TAC) for Northeast Arctic cod (NEA cod):
-estimate the average TAC level for the coming 3 years based on $\mathrm{F}_{\mathrm{pa}}$. TAC for the next year will be set to this level as a starting value for the 3-year period.
-the year after, the TAC calculation for the next 3 years is repeated based on the updated information about the stock development, however the TAC should not be changed by more than + - $10 \%$ compared with the previous year's TAC. If the TAC, by following such a rule, corresponds to a fishing mortality (F) lower than 0.30 the TAC should be increased to a level corresponding to a fishing mortality of 0.30 .
-if the spawning stock falls below $\mathrm{B}_{\mathrm{pa}}$, the procedure for establishing TAC should be based on a fishing mortality that is linearly reduced from $\mathrm{F}_{\mathrm{pa}}$ at $\mathrm{B}_{\mathrm{pa}}$, to $\mathrm{F}=0$ at SSB equal to zero. At SSB-levels below $\mathrm{B}_{\mathrm{p} a}$ in any of the operational years (current year, a year before and 3 years of prediction) there should be no limitations on the year-to-year variations in TAC."
ICES have been requested to evaluate whether this amended rule is in accordance with the precautionary approach (see section 0.3). The results of this evaluation are given in section 3.12.

### 3.6.4 Target reference points

The Russian-Norwegian Fishery Commission has requested an evaluation of the maximum sustainable yield (MSY) from the Barents Sea, taking into account species interactions and the influence from the environment. The work shall start with cod and gradually incorporate other species. A first step towards this is to study the MSY of cod in a single-species context (Kovalev and Bogstad, 2005). They studied the longterm yield of cod using the same biological model as used in the evaluation of the harvest control rule. Thus, mean weight at age in the stock was modelled as a function of total stock size, and mean weight at age in the catch and maturity at age was modelled as a function of mean weight at age in the stock. Cannibalism was included, and a stochastic segmented regression SSB-recruitment relationship was used. The results indicated that the long-term yield is fairly stable for a range of fishing mortalities between 0.25 and 0.6 . Density dependent effects in cannibalism and growth are considered as the main reasons for this rather wide F-range with stable high yield. It should be noted that there are few observations of biological parameters for low fishing mortalities and high stock sizes, so that the results for low Fs are more uncertain than those for higher Fs.

### 3.7 Prediction (Table 3.27-3.29)

### 3.7.1 Prediction input (Tables 3.27, Figure 3.10a-b, 3.11)

The input data to the short-term prediction with management option table (20102013) are given in Table 3.27. For 2010 stock weights and maturity were taken from surveys as described in Sections 3.32 and 3.3.4.
Catch weights in 2010 onwards and stock weights in 2011 onwards are predicted by the method described by Brander (2002), where the latest observation of weights by cohort are used together with average annual increments to predict the weight of the cohort the following year.
$W(a+1, y+1)=W(a, y)+\operatorname{Incr}(a)$, where $\operatorname{Incr}(a)$ is a "medium term" average of $\operatorname{Incr}(a, y)=$ $W(a+1, y+1)-W(a, y)$
This method was introduced in the cod prediction in the 2003 w orking group. Then it was decided that for Catch Weights average annual increments by age were calculated for the period 1994-2001, and for Stock Weights average annual increments by
age were calculated for the period 1995-2002. At the 2004 working group it was decided to follow the same procedure, except that for stock weights the period (20012003) was chosen for calculating average annual increment. The reason was that those years indicate a declining trend that could be associated with declining capelin stock. The same argument was considered valid at the 2005 and later working groups and only the 3 most recent values of annual increments were used for predicting stock weights. For catch weights, we use a 10-year period (2000-2009) for averaging the increments. Figures 3.10a and 3.10b show how these predictions perform back in history.

The maturity ogive for the years 2011 and 2012 was predicted by using the 2008-2010 average. The exploitation pattern in 2010 and later years was set equal to the 20072009 average.

The stock number at age in 2010 was taken from the final VPA (Table3.22) for ages 4 and older. The recruitment at age 3 in the years 2010-2012 was estimated as described in section 1.4.5. Figure 3.11 shows the development in natural mortality due to cannibalism for cod (prey) age groups 1-3 together with the abundance of capelin in the period 1984-2009. The recent 3 years average $M$ was used as input for the years 20102012 in the prediction.

For 2013, the 2012 values were used for all input data, except for recruitment, where the long-term arithmetic mean ( 600 million at age 3 ) was used.

The assessment shows a decrease in F from 2006 to 2009. Effort has also decreased (Figure 3.15), and thus similar to last year's assessment F in 2009 is considered to be a better estimate for F in the intermediate year (2010) than the estimate using three year average F. Table 3.27 shows input data to the predictions.

### 3.7.2 Prediction results (Tables 3.28, 3.29b)

The catches corresponding to $\mathrm{Fsq}_{\mathrm{sq}}$ in 2010 is 593000 tonnes (Table 3.28). This is close to the TAC for 2010 ( 607000 tonnes). The resulting SSB in 2011 is $1,488,000$ tonnes. Table 3.28 also shows the short-term consequences over a range of F-values in 2011. The detailed outputs corresponding to $\mathrm{F}_{\text {sq }}$ in 2010, the F corresponding to the HCR in 2011 and $\mathrm{F}_{\mathrm{pa}}$ in 2012-2013 is given in Table 3.29b. Summarised results are shown in text table below.

| Rationale | Landings ${ }^{1)}$ <br> $(2011)$ | Basis | F <br> $(2011)$ | SSB <br> $(2012)$ | \%SSB <br> change ${ }^{2)}$ | \% TAC <br> change |
| :--- | :---: | :--- | :--- | :--- | :---: | :---: |
| Zero catch | 0 | $0^{*} \mathrm{~F}_{\mathrm{sq}}$ | 0 | 2192 | +47 | -100 |
| Agreed <br> mana ge ment <br> Plan $^{4}$ | 703 | $1.09^{*} \mathrm{~F}_{\mathrm{sq}}$ | 0.30 | 1689 | +14 | +16 |
| Status quo | 654 | $1.00^{*} \mathrm{~F}_{\mathrm{sq}}$ | 0.28 | 1731 | +16 | +8 |
| Precautionary <br> Limits | 896 | $\mathrm{~F}_{\mathrm{pa}}$ | 0.40 | 1527 | +3 | +48 |

Weights in ' 000 t .
${ }^{1)}$ Landings are total landings without IUU landings. If this figure is taken as T AC, no implementation error is assumed.
${ }^{2)}$ SSB 2012 relative to SSB 2011.
${ }^{3)}$ T AC 2011 relative to T AC 2010.
${ }^{4)}$ Forecast based on $\mathrm{F}=\mathbf{0 . 3 0}$.

This catch forecast covers all catches. It is then implied that all types of catches are to be counted against this TAC. It also means that if any overfishing is expected to take place, the above calculated TAC should be reduced by the expected amount of overfishing.

### 3.8 Comparison with last year's assessment

The text table below compares this year's estimates with last year's estimates for the year 2009 numbers at age (millions), total biomass, spawning biomass (thousand tonnes), as well as reference F for the year 2008.


The final assessment values for ages 3,4,5,7 and 8 are fairly close (within $10 \%$ ) of the 2009 assessment, while age 6 have this year been revised upward by $17 \%$ and ages 9 and 10 downwards by 11 and $13 \%$, respectively. The F in 2008 is 0.03 ( $10 \%$ ) higher last year's estimate, but the total stock biomass and SSB in 2009 are very close to the previous estimates.

### 3.9 Additional assessment methods

### 3.9.1 Survey calibration method (Figures 3.12-13)

A "calibrated" prediction method of stock numbers from the Joint bottom trawl survey against VPA numbers, using data from the period 1981-1995 to scale the survey series to absolute numbers, was carried out. The method is described in Pennington and Nakken (WD14, 2008). The regression is done for ages $4-6$ and $7+$ separately. The results, using a regression method with intercept, are shown in Figures 3.12-3.13 The method compares well to the VPA results for stock abundance in 2010: Ages 4-6: Calibration method 1026 millions, vs. 1137 millions from VPA. Age 7+: Calibration method 284 millions, vs. 266 millions from VPA. The figures show a shift both for ages 46 and $7+$ occurring around 2006 for the relation between the survey calibration and theVPA.

### 3.9.2 Gadget (Figure 3.14)

Thebiological Gadget model used for Northeast Arctic cod is described in Bogstad et al. (2004). The same model as last year was run, updated with an additional year of data. Model runs are now performed using Gadget version 2.1.06. The trends obtained last year are also seen this year, with continuing increases in overall and spawning stock biomass, but low to moderate recruitment (Figure 3.14). The modeled historical stock is very similar to that from the previous year, with very slight upwards revisions in some years, mostly in the modeled SSB. The Gadget model is in broad agreement with the XSA model in that that current stock is close to the highest values seen over the last 20 years. There is some indication in the model results that recruitment may now be dropping from the recent high levels.

### 3.10 Comments to the assessment

The magnitude of IUU catches has decreased considerably from around $30 \%$ of official landings to $3 \%$ in 2008. No any IUU catches were registered for 2009. The uncertainty relating to total catch for the years 2002-2006 could still have significant influence on the assessment of the current stock.

XSA has for several years been used for the assessment of cod, but in recent years additional assessment models have been tried, e.g. the "survey calibration model", "Gadget", and "TISVPA". These models have given results characterized by differences in level of stock size and exploitation, although the trends have in most cases been similar.

The WG realizes that imprecise input data, in particular the catch-at-age matrix, could be a main obstacle to producing precise stock assessments, irrelevant of which model is used. The WG, therefore, recognizes the need for improvements to the input data, and in particular more reliable catch data (see chapter 0.5).

However, the WG also recognizes the need for a more thorough comparison of assessment methodologies. In particular, the models XSA and TISVPA would be interesting to explore and compare. These two models are related to the same class of cohort analysis models.

XSA model is used in many years by AFWG, that has a big experience to work with this model, but TISVPA has some advantages. In particular, TISVPA allows strictly to formulate a statistical meaning of solution, is more robust and reliable.

Benchmarking of various assessment models is not a trivial task, since criteria for performance are not easy to establish across models. Therefore, some guidance for how to perform such comparisons would be valued. It is also clear that a benchmark workshop should not be planned too early, since most of the work in connection with the benchmarking will have to be done prior to the workshop.

### 3.11 New data sources

This section describes some data sources, which could be included in the assessment in the future.

### 3.11.1 Catch data (Tables 3.30, 3.31, 3.1b)

Discard and bycatch data series (Table 3.30, 3.31) should be updated and then included in the catch at age matrix. Table 3.31 (taken from Ajiad et al., WD2, 2008) presents by-catch in the Norwegian shrimp fishery by cod age (previously this has been given by cod length). The by-catch mainly consists of age 1 and 2 fish, but the bycatch is generally small compared to other reported sources of mortality: catches, discards and the number of cod eaten by cod. From 1992 onwards, by-catches of age 3 and older fish are negligible, because use of sorting grids was made mandatory. However, in 1985, by-catches of age 5 and 6 cod were about one third of the reported catches for those age groups. The year class for which the by-catches were highest, was the 1983 year class (total by-catch of age 2 and older fish of about 60 million, compared to a stock estimate of about 1000 million at age 3 ).

Also the time series described by Hylen (2002), extending the VPA back to 1932, should be reviewed. Consistency between the catch data used for NEA cod and coastal cod should also be ensured. At present, the catch figures used in the coastal
cod assessment are not equal to the difference between the total cod catch and the catch used in the NEA cod assessment (Table 3.1b).

It could also be considered to take the difference in age at maturation determined by contemporary and historic age readers (Section 0.5) into account.

Updating the catch data series as indicated here will affect the reference points, but only to a small extent estimate of present stock size. These updates should all be carried out at the same time.

### 3.11.2 Consumption data

Work on extending the cannibalism time series back to 1947 is ongoing (Yaragina et al.2009a).

### 3.11.3 Survey data (Tables 3.14, A14)

The bottom trawl estimates from the joint ecosystem survey in August-September, starting in 2004. This survey covers the entire distribution area of cod. The new index values for period 2004-2009 become available for AFWG this year (Table A14, WD 20). This time series have been tested as new tuning fleet in XSA (Fleet 007). The single fleet estimates of survivors of all ages, based on this index were slightly lower compare to other single fleet runs (Table 3.14). The results of XSA run including all fleets (as in Final XSA + Fleet 007) were very close to final XSA ones. Analysis of XSA diagnostic from this run (WD 20) demonstrate that fleet 007 has reasonably good quality comparable to best time series (fleet 15 and 18). This index could be considered for use as a tuning series on next benchmark.

### 3.11.4 New CPUE series

The new biomass indices described in WD11 (2008) and 21 (2008), based on vessels' daily reports, may in the future be included in the tuning of assessment models.

### 3.12 Evaluation of amended HCR

### 3.12.1 Introduction

The harvest control rule for NEA cod was amended at the $38^{\text {th }}$ session of The Joint Norw egian-Russian Fishery Commission (see section 3.6.3.) and ICES were requested to evaluate the new rule.

The previous version of the HCR was evaluated by ICES in 2005. The long-term stochastic simulations were done by means of software (PROST) developed for the purpose of evaluating HCRs of this type (Åsnes, WD 2 to AFWG 2005). The same population model as that described in detail in the AFWG 2005 report was used this year for HCR evaluation (ICES 2005).

### 3.12.2 Mathematical formulation of the amended HCR

Let y denote the year for which the quota is to be set. Let the term 3-year rule (F1, x) denote applying the 3-year average rule described above with $\mathrm{F}_{5-10}=\mathrm{F} 1$ and an $\mathrm{x} \%$ limit on year-to-year changes in TAC. The limit on increase of TAC from year to year could be set different from the limit on decrease from year to year, but such asymmetric rules were not tested. It is assumed that $\operatorname{SSB}(\mathrm{y})$ is not affected by $\mathrm{F}(\mathrm{y})$, which is in line with the current settings used by AFWG (the proportion of F and M before spawning is set to 0 ). The rule can then be described in the following way:

$$
\begin{aligned}
& \text { If } \mathrm{SSB}(\mathrm{y})>\mathrm{B}_{\mathrm{pa}} \text { then } \\
& \qquad \begin{array}{l}
\text { if } \mathrm{SSB}(\mathrm{y}-1)>\mathrm{B}_{\mathrm{pa}} \text { and } \mathrm{SSB}(\mathrm{y}+1)>\mathrm{B}_{\mathrm{pa}} \text { and } \mathrm{SSB}(\mathrm{y}+2)>\mathrm{B}_{\mathrm{pa}}: \\
\qquad \mathrm{F}(\mathrm{y}) \text { set by 3-year rule }(0.40,10) \\
\text { if } \mathrm{F}(\mathrm{y})<0.30, \mathrm{~F}(\mathrm{y})=0.30
\end{array}
\end{aligned}
$$

else
$\mathrm{F}(\mathrm{y})$ set by 3-year rule( 0.40 , unconstrained)
else
$\mathrm{F}(\mathrm{y})$ set by 3-year rule $\left(0.40 * \operatorname{SSB}(\mathrm{y}) / \mathrm{B}_{\mathrm{pa}}\right.$, unconstrained $)$

### 3.12.3 Evaluation undertaken in 2010

In this evaluation, we choose to repeat the evaluation undertaken in 2005, as far as the biological model, settings, and input data are concerned. Values for distortion on input stock numbers etc. can be found in ICES (2005). The simulation tool (PROST) was however, changed to account for the change in the HCR. Like in 2005, we chose to do two runs with long-term simulation. The settings and the results are described in the text tables below.

| Run No. | Target F | M ages 3 and 4 (high: 0.7\&0.4, Low: 0.2\&0.2) |
| :---: | :---: | :---: |
| 1 | 0.40 | Low |
| 2 | 0.40 | High |


| Run <br> No. | Realized <br> F | Catch | TSB | SSB | Recruits | \% years <br> $\mathrm{SSB}<\mathrm{B}_{\mathrm{lim}}$ | \% years <br> $\mathrm{SSB}<\mathrm{B}_{\mathrm{pa}}$ | Average <br> year-to-year <br> \% change in <br> inAC | \% years <br> where F 0.3 <br> part of rule <br> is active |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.64 | 896 | 3036 | 566 | 687 | 0.00 | 14.95 | 15 | 3 |
| 2 | 0.60 | 477 | 1813 | 430 | 684 | 0.18 | 52.57 | 19 | 0 |

If the result table is compared to table 3.36 in ICES (2005), it is seen that the differences in catch, TSB, SSB and recruits are small (mostly within 5\%) and the overall results are the same; namely that the number of years in the simulation period where the SSB drops below Blim, is negligible in either case.

The software keeps track of which part of the rule is active in each run. The new part of the rule (if $\mathrm{F}(\mathrm{y})<0.30, \mathrm{~F}(\mathrm{y})=0.30$ ) is active in about $3 \%$ of the years in run no. 1, while it is not active at all in run no. 2. The average year-to-year change in TAC is approximately the same for both runs.

It is concluded that the change made to the rule by the MRNC in 2009 does not affect the outcome of long-term stochastic simulation runs to any noticeable degree. Consequently, the amended HCR for NEA cod could be characterised as being in accordance with the precautionary approach to fisheries management.

### 3.13 Answering 2009 comments from Reviewers:

The minutes of the review of the 2009 AFWG report contained a number of comments to the NEA cod assessment. Below is a summary how AFWG has responded to this:

No discussion regarding comparisons of TISVPA and XSA models took place at the current meeting because the Group was dispersed around Europe (see Introduction section). Many reviewers' comments are related to this problem.

Comment regarding not inclusion information about discarding was taken into account and appropriate sentences were included in the report.

The other comments need to be considered during the next benchmark meeting.

Table 3.1a North-East Arctic COD. Total catch (t) by fishing areas and unreported catch. (Data provided by Working Group members.)

| Year | Sub-area I | Division lla | Division llb | Unreported catches | Total catch |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 409,694 | 153,019 | 220,508 |  | 783,221 |
| 1962 | 548,621 | 139,848 | 220,797 |  | 909,266 |
| 1963 | 547,469 | 117,100 | 111,768 |  | 776,337 |
| 1964 | 206,883 | 104,698 | 126,114 |  | 437,695 |
| 1965 | 241,489 | 100,011 | 103,430 |  | 444,983 |
| 1966 | 292,253 | 134,805 | 56,653 |  | 483,711 |
| 1967 | 322,798 | 128,747 | 121,060 |  | 572,605 |
| 1968 | 642,452 | 162,472 | 269,254 |  | 1,074,084 |
| 1969 | 679,373 | 255,599 | 262,254 |  | 1,197,226 |
| 1970 | 603,855 | 243,835 | 85,556 |  | 933,246 |
| 1971 | 312,505 | 319,623 | 56,920 |  | 689,048 |
| 1972 | 197,015 | 335,257 | 32,982 |  | 565,254 |
| 1973 | 492,716 | 211,762 | 88,207 |  | 792,685 |
| 1974 | 723,489 | 124,214 | 254,730 |  | 1,102,433 |
| 1975 | 561,701 | 120,276 | 147,400 |  | 829,377 |
| 1976 | 526,685 | 237,245 | 103,533 |  | 867,463 |
| 1977 | 538,231 | 257,073 | 109,997 |  | 905,301 |
| 1978 | 418,265 | 263,157 | 17,293 |  | 698,715 |
| 1979 | 195,166 | 235,449 | 9,923 |  | 440,538 |
| 1980 | 168,671 | 199,313 | 12,450 |  | 380,434 |
| 1981 | 137,033 | 245,167 | 16,837 |  | 399,037 |
| 1982 | 96,576 | 236,125 | 31,029 |  | 363,730 |
| 1983 | 64,803 | 200,279 | 24,910 |  | 289,992 |
| 1984 | 54,317 | 197,573 | 25,761 |  | 277,651 |
| 1985 | 112,605 | 173,559 | 21,756 |  | 307,920 |
| 1986 | 157,631 | 202,688 | 69,794 |  | 430,113 |
| 1987 | 146,106 | 245,387 | 131,578 |  | 523,071 |
| 1988 | 166,649 | 209,930 | 58,360 |  | 434,939 |
| 1989 | 164,512 | 149,360 | 18,609 |  | 332,481 |
| 1990 | 62,272 | 99,465 | 25,263 | 25,000 | 212,000 |
| 1991 | 70,970 | 156,966 | 41,222 | 50,000 | 319,158 |
| 1992 | 124,219 | 172,532 | 86,483 | 130,000 | 513,234 |
| 1993 | 195,771 | 269,383 | 66,457 | 50,000 | 581,611 |
| 1994 | 353,425 | 306,417 | 86,244 | 25,000 | 771,086 |
| 1995 | 251,448 | 317,585 | 170,966 |  | 739,999 |
| 1996 | 278,364 | 297,237 | 156,627 |  | 732,228 |
| 1997 | 273,376 | 326,689 | 162,338 |  | 762,403 |
| 1998 | 250,815 | 257,398 | 84,411 |  | 592,624 |
| 1999 | 159,021 | 216,898 | 108,991 |  | 484,910 |
| 2000 | 137,197 | 204,167 | 73,506 |  | 414,870 |
| 2001 | 142,628 | 185,890 | 97,953 |  | 426,471 |
| $2002{ }^{\text {² }}$ | 184,789 | 189,013 | 71,242 | 90000/21716 | 535045/466760 |
| $2003{ }^{\text {F }}$ | 163,109 | 222,052 | 51,829 | 115000/27748 | 551990/464738 |
| $2004{ }^{\text {F2 }}$ | 177,888 | 219,261 | 92,296 | 117000/30000 | 606445/519445 |
| $2005{ }^{\text {52 }}$ | 159,573 | 194,644 | 121,059 | 166000/41000 | 641276/516276 |
| $2006{ }^{\text {F }}$ | 159,851 | 204,603 | 104,743 | 127000/28000 | 596197/497197 |
| $2007{ }^{\text {F }}$ | 152,522 | 195,383 | 97,891 | 41087/8757 | 486883/454553 |
| 2008 | 144,905 | 203,244 | 101,022 | 15000/0 | 464171/449171 |
| $2009{ }^{\text {¹ }}$ | 161,602 | 207,205 | 154,623 |  | 523,431 |

${ }^{1}$ Provisional figures.
${ }^{2}$ two alternative estimates (see Chapter 3.1.3 of the 2008 AFWG Report for further details)

Table 3.1b Landings of Norwegian Coastal Cod in Sub-areas I and II

| Year | Landings in '000 t |  |
| :---: | :---: | :---: |
|  | As calculated from samples and reported to AFWG | By area and time of capture |
| 1960 | - | 43 |
| 1961 | - | 32 |
| 1962 | - | 30 |
| 1963 | - | 40 |
| 1964 | - | 46 |
| 1965 | - | 24 |
| 1966 | - | 29 |
| 1967 | - | 33 |
| 1968 | - | 47 |
| 1969 | - | 52 |
| 1970 | - | 49 |
| 1971 | - | *) |
| 1972 | - | *) |
| 1973 | - | *) |
| 1974 | - | *) |
| 1975 | - | *) |
| 1976 | - | *) |
| 1977 | - | *) |
| 1978 | - | *) |
| 1979 | - | *) |
| 1980 | - | 40 |
| 1981 | - | 49 |
| 1982 | - | 42 |
| 1983 | - | 38 |
| 1984 | 74 | 33 |
| 1985 | 75 | 28 |
| 1986 | 69 | 26 |
| 1987 | 61 | 31 |
| 1988 | 59 | 22 |
| 1989 | 40 | 17 |
| 1990 | 28 | 24 |
| 1991 | 25 | 25 |
| 1992 | 42 | 35 |
| 1993 | 53 | 44 |
| 1994 | 55 | 48 |
| 1995 | 57 | 39 |
| 1996 | 62 | 32 |
| 1997 | 63 | 36 |
| 1998 | 52 | 29 |
| 1999 | 41 | 23 |
| 2000 | 37 | 19 |
| 2001 | 30 | 14 |
| 2002 | 41 | 20 |
| 2003 | 35 | 19 |
| 2004 | 25 | 14 |
| 2005 | 22 | 13 |
| 2006 | 26 | 15 |
| 2007 | 24 | 13 |
| 2008** | 26 | 13 |
| 2009 | 25 | 15 |
| Average 1984-2009 | 44 | - 25 |
| *) No data |  |  |
| ** Corrected |  |  |

Table 3.2 North-East Arctic COD. Total nominal catch ('OOO t) by trawl and other gear for each area, data provided by Working Group members

|  | Sub-area I |  | Division Ila |  | Division llb |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Trawl | Others | Trawl | Others | Trawl | Others |
| 1967 | 238.0 | 84.8 | 38.7 | 90.0 | 121.1 | - |
| 1968 | 588.1 | 54.4 | 44.2 | 118.3 | 269.2 | - |
| 1969 | 633.5 | 45.9 | 119.7 | 135.9 | 262.3 | - |
| 1970 | 524.5 | 79.4 | 90.5 | 153.3 | 85.6 | - |
| 1971 | 253.1 | 59.4 | 74.5 | 245.1 | 56.9 | - |
| 1972 | 158.1 | 38.9 | 49.9 | 285.4 | 33.0 | - |
| 1973 | 459.0 | 33.7 | 39.4 | 172.4 | 88.2 | - |
| 1974 | 677.0 | 46.5 | 41.0 | 83.2 | 254.7 | - |
| 1975 | 526.3 | 35.4 | 33.7 | 86.6 | 147.4 | - |
| 1976 | 466.5 | 60.2 | 112.3 | 124.9 | 103.5 | - |
| 1977 | 471.5 | 66.7 | 100.9 | 156.2 | 110.0 | - |
| 1978 | 360.4 | 57.9 | 117.0 | 146.2 | 17.3 | - |
| 1979 | 161.5 | 33.7 | 114.9 | 120.5 | 8.1 | - |
| 1980 | 133.3 | 35.4 | 83.7 | 115.6 | 12.5 | - |
| 1981 | 91.5 | 45.1 | 77.2 | 167.9 | 17.2 | - |
| 1982 | 44.8 | 51.8 | 65.1 | 171.0 | 21.0 | - |
| 1983 | 36.6 | 28.2 | 56.6 | 143.7 | 24.9 | - |
| 1984 | 24.5 | 29.8 | 46.9 | 150.7 | 25.6 | - |
| 1985 | 72.4 | 40.2 | 60.7 | 112.8 | 21.5 | - |
| 1986 | 109.5 | 48.1 | 116.3 | 86.4 | 69.8 | - |
| 1987 | 126.3 | 19.8 | 167.9 | 77.5 | 129.9 | 1.7 |
| 1988 | 149.1 | 17.6 | 122.0 | 88.0 | 58.2 | 0.2 |
| 1989 | 144.4 | 19.5 | 68.9 | 81.2 | 19.1 | 0.1 |
| 1990 | 51.4 | 10.9 | 47.4 | 52.1 | 24.5 | 0.8 |
| 1991 | 58.9 | 12.1 | 73.0 | 84.0 | 40.0 | 1.2 |
| 1992 | 103.7 | 20.5 | 79.7 | 92.8 | 85.6 | 0.9 |
| 1993 | 165.1 | 30.7 | 155.5 | 113.9 | 66.3 | 0.2 |
| 1994 | 312.1 | 41.3 | 165.8 | 140.6 | 84.3 | 1.9 |
| 1995 | 218.1 | 33.3 | 174.3 | 143.3 | 160.3 | 10.7 |
| 1996 | 248.9 | 32.7 | 137.1 | 159.0 | 147.7 | 6.8 |
| 1997 | 235.6 | 37.7 | 150.5 | 176.2 | 154.7 | 7.6 |
| 1998 | 219.8 | 31.0 | 127.0 | 130.4 | 82.7 | 1.7 |
| 1999 | 133.3 | 25.7 | 101.9 | 115.0 | 107.2 | 1.8 |
| 2000 | 111.7 | 25.5 | 105.4 | 98.8 | 72.2 | 1.3 |
| 2001 | 119.1 | 23.5 | 83.1 | 102.8 | 95.4 | 2.5 |
| 2002 | 147.4 | 37.4 | 83.4 | 105.6 | 69.9 | 1.3 |
| 2003 | 146.0 | 17.1 | 107.8 | 114.2 | 50.1 | 1.8 |
| 2004 | 154.4 | 23.5 | 100.3 | 118.9 | 88.8 | 3.5 |
| 2005 | 132.4 | 27.2 | 87.0 | 107.7 | 115.4 | 5.6 |
| 2006 | 141.8 | 18.1 | 91.2 | 113.4 | 100.1 | 4.6 |
| 2007 | 129.6 | 22.9 | 84.8 | 110.6 | 91.6 | 6.3 |
| 2008 | 123.8 | 21.1 | 94.8 | 108.4 | 95.3 | 5.7 |
| $2009{ }^{\text {¹ }}$ | 130.1 | 31.5 | 102.0 | 105.2 | 142.1 | 11.4 |
| 1 Provis | ional figu | ures. |  |  |  |  |

Table 3.3 North-East Arctic COD. Nominal catch (t) by countries
(Sub-area I and Divisions lla and llb combined, data provided by Working Group members.)



Table 3.5 North-East Arctic COD. Basis for maturity ogives (percent) used in the assessment. Norwegian and Russian data.

| Norway |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :---: |
| Percentage mature |  |  |  |  |  |  |  |  |  |
| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |  |
|  |  |  |  |  |  |  |  |  |  |
| 1982 | - | 5 | 10 | 34 | 65 | 82 | 92 | 100 |  |
| 1983 | 5 | 8 | 10 | 30 | 73 | 88 | 97 | 100 |  |

Russia

|  | Percentage mature <br> Age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1984 | - | 5 | 18 | 31 | 56 | 90 | 99 | 100 |
| 1985 |  | 1 | 10 | 33 | 59 | 85 | 92 | 100 |
| 1986 |  | 2 | 9 | 19 | 56 | 76 | 89 | 100 |
| 1987 |  | 1 | 9 | 23 | 27 | 61 | 81 | 80 |
| 1988 |  | 1 | 3 | 25 | 53 | 79 | 100 | 100 |
| 1989 | 0.0 | 0.0 | 2.0 | 15.0 | 39.0 | 59.0 | 83.0 | 100.0 |
| 1990 | 0.0 | 2.0 | 6.0 | 20.0 | 47.0 | 62.0 | 81.0 | 95.0 |
| 1991 | 0.0 | 3.0 | 1.0 | 23.0 | 66.0 | 82.0 | 96.0 | 100.0 |
| 1992 | 0.0 | 1.0 | 8.0 | 31.0 | 73.0 | 92.0 | 95.0 | 100.0 |
| 1993 | 0.0 | 3.0 | 7.0 | 21.0 | 56.0 | 89.0 | 95.0 | 99.0 |
| 1994 | 0.0 | 1.0 | 8.0 | 30.0 | 55.0 | 84.0 | 95.0 | 98.0 |
| 1995 | 0.0 | 0.0 | 4.0 | 23.0 | 61.0 | 75.0 | 94.0 | 97.0 |
| 1996 | 0.0 | 0.0 | 1.0 | 22.0 | 56.0 | 82.0 | 95.0 | 100.0 |
| 1997 | 0.0 | 0.0 | 1.0 | 10.0 | 48.0 | 73.0 | 90.0 | 100.0 |
| 1998 | 0.0 | 0.0 | 2.0 | 15.0 | 47.0 | 87.0 | 97.0 | 96.0 |
| 1999 | 0.0 | 0.2 | 1.3 | 9.9 | 38.4 | 74.9 | 94.0 | 100.0 |
| 2000 | 0.0 | 0.0 | 6.0 | 19.2 | 51.4 | 84.0 | 95.5 | 100.0 |
| 2001 | 0.1 | 0.1 | 3.9 | 27.9 | 62.3 | 89.4 | 96.3 | 100.0 |
| 2002 | 0.1 | 1.9 | 10.9 | 34.4 | 68.1 | 82.8 | 97.6 | 100.0 |
| 2003 | 0.2 | 0.0 | 11.0 | 29.2 | 65.9 | 89.6 | 95.1 | 100.0 |
| 2004 | 0.0 | 0.7 | 8.0 | 33.8 | 63.3 | 83.4 | 96.4 | 96.4 |
| 2005 | 0.0 | 0.6 | 4.6 | 24.2 | 61.5 | 84.9 | 95.3 | 98.1 |
| 2006 | 0.0 | 0.0 | 6.1 | 29.6 | 59.6 | 89.5 | 96.4 | 100.0 |
| 2007 | 0.0 | 0.4 | 5.7 | 20.8 | 60.4 | 83.5 | 96.0 | 100.0 |
| 2008 | 0.0 | 0.5 | 4.0 | 24.6 | 48.3 | 84.4 | 94.7 | 98.7 |
| 2009 | 0.0 | 0.0 | 6.0 | 28.0 | 66.0 | 85.0 | 97.0 | 100.0 |
| 2010 | 0.0 | 0.2 | 1.5 | 22.8 | 47.0 | 77.4 | 90.2 | 95.5 |
| Norway |  |  |  |  |  |  |  |  |

Percentage mature

| Age |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1985 | - | 1 | 9 | 38 | 51 | 85 | 100 | 79 |
| 1986 | 3 | 7 | 8 | 19 | 50 | 67 | 36 | 80 |
| 1987 | - | 0 | 4 | 12 | 16 | 31 | 19 | - |
| 1088 |  | 2 | 0 | 41 | 54 | 45 | 100 | 100 |

$\begin{array}{llllllll}1988 & - & 2 & 6 & 41 & 54 & 45 & 100\end{array} 100$
$\begin{array}{lllllllll}1989 & 1.5 & 0.7 & 3.9 & 30.7 & 70.4 & 82.0 & 100.0 & 100.0\end{array}$
$\begin{array}{lllllllll}1990 & 1.5 & 0.7 & 4.2 & 22.0 & 57.5 & 80.9 & 100.0 & 100.0\end{array}$
$\begin{array}{lllllllll}1991 & 0.1 & 3.4 & 13.9 & 38.0 & 75.5 & 90.1 & 95.4 & 100.0\end{array}$
$\begin{array}{lllllllll}1992 & 0.2 & 1.9 & 21.0 & 52.8 & 87.0 & 96.5 & 99.8 & 100.0\end{array}$
$\begin{array}{lllllllll}1993 & 0.0 & 2.6 & 10.4 & 52.6 & 84.8 & 97.2 & 99.3 & 99.7\end{array}$
$\begin{array}{lllllllll}1994 & 0.5 & 0.3 & 15.8 & 36.9 & 62.8 & 88.4 & 97.6 & 100.0\end{array}$
$\begin{array}{lllllllll}1995 & 0.0 & 0.6 & 8.2 & 51.5 & 63.8 & 81.1 & 98.0 & 99.3\end{array}$
$\begin{array}{lllllllll}1996 & 0.0 & 0.0 & 2.8 & 29.6 & 70.2 & 82.1 & 100.0 & 100.0\end{array}$
$\begin{array}{lllllllll}1997 & 0.0 & 0.0 & 1.5 & 17.9 & 73.3 & 93.0 & 99.1 & 100.0\end{array}$
$\begin{array}{lllllllll}1998 & 0.1 & 0.7 & 3.2 & 15.4 & 47.3 & 75.7 & 94.3 & 100.0\end{array}$
$\begin{array}{lllllllll}1999 & 0.4 & 0.2 & 1.6 & 27.5 & 70.5 & 94.6 & 99.0 & 100.0\end{array}$
$\begin{array}{lllllllll}2000 & 0.0 & 0.1 & 8.2 & 30.2 & 77.3 & 81.9 & 100.0 & 100.0\end{array}$
$\begin{array}{lllllllll}2001 & 0.5 & 0.5 & 9.0 & 43.8 & 62.5 & 74.4 & 94.1 & 100.0\end{array}$
$\begin{array}{lllllllll}2002 & 0.3 & 0.7 & 5.9 & 43.2 & 68.4 & 85.3 & 92.5 & 100.0\end{array}$
$\begin{array}{lllllllll}2003 & 0.0 & 0.2 & 6.5 & 36.0 & 68.6 & 88.0 & 96.3 & 100.0\end{array}$
$\begin{array}{lllllllll}2004 & 0.2 & 1.4 & 10.2 & 54.6 & 81.8 & 90.9 & 98.8 & 98.9\end{array}$
$\begin{array}{lllllllll}2005 & 0.0 & 0.3 & 9.0 & 55.2 & 81.8 & 93.5 & 98.0 & 100.0\end{array}$
$\begin{array}{lllllllll}2006 & 0.0 & 0.2 & 5.9 & 44.3 & 69.8 & 89.9 & 96.7 & 100.0\end{array}$
$\begin{array}{lllllllll}2007 & 0.1 & 0.3 & 8.7 & 47.9 & 84.3 & 91.7 & 99.1 & 100.0\end{array}$
$\begin{array}{lllllllll}2008 & 0.0 & 0.3 & 8.4 & 31.8 & 59.3 & 88.2 & 90.9 & 100.0\end{array}$
$\begin{array}{lllllllll}2009 & 0.0 & 0.0 & 9.2 & 46.3 & 85.0 & 86.4 & 98.4 & 99.3\end{array}$
$\begin{array}{lllllllll}2010 & 0.0 & 0.4 & 7.5 & 41.8 & 67.7 & 90.1 & 95.3 & 98.6\end{array}$

Table 3.6. Northeast artic cod. Barents Sea winter survey. Area covered ('000 square nautical miles) and areas implied in the method used to adjust for missing coverage in Russian Economic Zone. In 4 of the 5 adjusted years the adjustments were not based on area ratios, but the "index ratio by age" was used. This means that the index by age (for the area outside REZ) was scaled by the observed ratio between total index and the index outside REZ observed in the years prior to the survey.

| Year | Area covered | Additional area <br> implied in <br> adjustment | Adjustment method |
| :--- | :--- | :--- | :--- |
| $1981-92$ | 88.1 |  |  |
| 1993 | 137.6 |  |  |
| 1994 | 143.8 |  |  |
| 1995 | 186.6 |  | Index ratio |
| 1996 | 165.3 |  | Index ratio |
| 1997 | 87.5 | 78.0 |  |
| 1998 | 99.2 | 78.0 |  |
| 1999 | 118.3 |  |  |
| 2000 | 162.4 |  |  |
| 2001 | 164.1 |  |  |
| 2002 | 156.7 |  | Partly covered strata raised to full strata area |
| 2003 | 146.6 |  |  |
| 2004 | 164.6 |  |  |
| 2005 | 178.9 |  |  |
| 2006 | 169.1 | 18.1 |  |
| 2007 | 122.2 | 56.7 |  |
| 2008 | 164.4 |  |  |
| 2009 | 170.9 |  |  |
| 2010 | 159.9 |  |  |

Table 3.7
NE Arctic cod. International catch (thousands) at age for ages 1-15+


Table 3.8. Northeast arctic cod. Total number of cod (million) consumed by cod, by year and prey age group


[^2]Table 3.9. North-East Arctic COD. Catch numbers at age

Run title : Arctic Cod (run: SVPASA15/V15)
At 22/04/2010 19:53

|  | Table 1 | Catch numbers at age |  |  |  | Numbers*10**-3 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAF | 1946 | 1947 | 1948 | 1949 |  |  |  |  |  |  |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 4008 | 710 | 140 | 991 |  |  |  |  |  |  |
|  | 4 | 10387 | 13192 | 3872 | 6808 |  |  |  |  |  |  |
|  | 5 | 18906 | 43890 | 31054 | 35214 |  |  |  |  |  |  |
|  | 6 | 16596 | 52017 | 55983 | 100497 |  |  |  |  |  |  |
|  | 7 | 13843 | 45501 | 77375 | 83283 |  |  |  |  |  |  |
|  | 8 | 15370 | 13075 | 21482 | 29727 |  |  |  |  |  |  |
|  | 9 | 59845 | 19718 | 15237 | 13207 |  |  |  |  |  |  |
|  | 10 | 22618 | 47678 | 9815 | 5606 |  |  |  |  |  |  |
|  | 11 | 10093 | 31392 | 30041 | 8617 |  |  |  |  |  |  |
|  | 12 | 9573 | 9348 | 7945 | 13154 |  |  |  |  |  |  |
|  | +gp | 8137 | 18055 | 12595 | 7719 |  |  |  |  |  |  |
| 0 | TOTAL | 189376 | 294576 | 265539 | 304823 |  |  |  |  |  |  |
|  | TONSL | 706000 | 882017 | 774295 | 800122 |  |  |  |  |  |  |
|  | SOPCC | 103 | 91 | 89 | 99 |  |  |  |  |  |  |
|  | Table 1 | Catch numbers at age |  |  |  | Numbers*10**-3 |  |  |  |  |  |
|  | YEAF | 1950 | 1951 | 1952 | 1953 | 1954 | 1955 | 1956 | 1957 | 1958 | 1959 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 1281 | 24687 | 24099 | 47413 | 11473 | 3902 | 10614 | 17321 | 31219 | 32308 |
|  | 4 | 10954 | 77924 | 120704 | 107659 | 155171 | 37652 | 24172 | 33931 | 133576 | 77942 |
|  | 5 | 29045 | 64013 | 113203 | 112040 | 146395 | 201834 | 129803 | 27182 | 71051 | 148285 |
|  | 6 | 45233 | 46867 | 73827 | 55500 | 100751 | 161336 | 250472 | 70702 | 40737 | 53480 |
|  | 7 | 62579 | 37535 | 49389 | 22742 | 40635 | 84031 | 86784 | 87033 | 38380 | 18498 |
|  | 8 | 30037 | 33673 | 20562 | 16863 | 10713 | 30451 | 51091 | 39213 | 35786 | 17735 |
|  | 9 | 19481 | 23510 | 24367 | 10559 | 11791 | 13713 | 14987 | 17747 | 13338 | 23118 |
|  | 10 | 9172 | 10589 | 15651 | 10553 | 8557 | 9481 | 7465 | 6219 | 10475 | 9483 |
|  | 11 | 6019 | 4221 | 8327 | 5637 | 6751 | 4140 | 3952 | 3232 | 3289 | 3748 |
|  | 12 | 4133 | 1288 | 3565 | 1752 | 2370 | 2406 | 1655 | 1220 | 1070 | 997 |
|  | +gp | 9862 | 4935 | 2158 | 797 | 1287 | 1350 | 1906 | 819 | 433 | 513 |
| 0 | TOTAL | 227796 | 329242 | 455852 | 391515 | 495894 | 550296 | 582901 | 304619 | 379354 | 386107 |
|  | TONSL | 731982 | 827180 | 876795 | 695546 | 826021 | 1147841 | 1343068 | 792557 | 769313 | 744607 |
|  | SOPCC | 109 | 115 | 93 | 105 | 93 | 106 | 105 | 100 | 112 | 93 |
|  | 1 |  |  |  |  |  |  |  |  |  |  |
|  | Table 1 | Catch numbers at age |  |  |  | Numbers*10**-3 |  |  |  |  |  |
|  | YEAF | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 37882 | 45478 | 42416 | 13196 | 5298 | 15725 | 55937 | 34467 | 3709 | 2307 |
|  | 4 | 97865 | 132655 | 170566 | 106984 | 45912 | 25999 | 55644 | 160048 | 174585 | 24545 |
|  | 5 | 64222 | 123458 | 167241 | 205549 | 97950 | 78299 | 34676 | 69235 | 267961 | 238511 |
|  | 6 | 67425 | 51167 | 89460 | 95498 | 58575 | 68511 | 42539 | 22061 | 107051 | 181239 |
|  | 7 | 23117 | 38740 | 28297 | 35518 | 19642 | 25444 | 37169 | 26295 | 26701 | 79363 |
|  | 8 | 8429 | 17376 | 21996 | 16221 | 9162 | 8438 | 18500 | 25139 | 16399 | 26989 |
|  | 9 | 7240 | 5791 | 7956 | 11894 | 6196 | 3569 | 5077 | 11323 | 11597 | 13463 |
|  | 10 | 11675 | 6778 | 2728 | 3884 | 3553 | 1467 | 1495 | 2329 | 3657 | 5092 |
|  | 11 | 4504 | 5560 | 2603 | 1021 | 783 | 1161 | 380 | 687 | 657 | 1913 |
|  | 12 | 1843 | 1682 | 1647 | 1025 | 172 | 131 | 403 | 316 | 122 | 414 |
|  | +gp | 682 | 1298 | 775 | 784 | 782 | 337 | 156 | 279 | 240 | 190 |
| 0 | TOTAL | 324884 | 429983 | 535685 | 491574 | 248025 | 229081 | 251976 | 352179 | 612679 | 574026 |
|  | TONSL | 622042 | 783221 | 909266 | 776337 | 437695 | 444930 | 483711 | 572605 | 1074084 | 1197226 |
|  | SOPCC | 104 | 110 | 124 | 102 | 103 | 129 | 123 | 109 | 108 | 105 |
|  | Table 1 | Catch numbers at age |  |  |  | Numbers*10**-3 |  |  |  |  |  |
|  | YEAF | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 7164 | 7754 | 35536 | 294262 | 91855 | 45282 | 85337 | 39594 | 78822 | 8600 |
|  | 4 | 10792 | 13739 | 45431 | 131493 | 437377 | 59798 | 114341 | 168609 | 45400 | 77484 |
|  | 5 | 25813 | 11831 | 26832 | 61000 | 203772 | 226646 | 79993 | 136335 | 88495 | 43677 |
|  | 6 | 137829 | 9527 | 12089 | 20569 | 47006 | 118567 | 118236 | 52925 | 56823 | 31943 |
|  | 7 | 96420 | 59290 | 7918 | 7248 | 12630 | 29522 | 47872 | 61821 | 25407 | 16815 |
|  | 8 | 31920 | 52003 | 34885 | 8328 | 4370 | 9353 | 13962 | 23338 | 31821 | 8274 |
|  | 9 | 8933 | 12093 | 22315 | 19130 | 2523 | 2617 | 4051 | 5659 | 9408 | 10974 |
|  | 10 | 3249 | 2434 | 4572 | 4499 | 5607 | 1555 | 936 | 1521 | 1227 | 1785 |
|  | 11 | 1232 | 762 | 1215 | 677 | 2127 | 1928 | 558 | 610 | 913 | 427 |
|  | 12 | 260 | 418 | 353 | 195 | 322 | 575 | 442 | 271 | 446 | 103 |
|  | +gp | 180 | 216 | 476 | 195 | 296 | 283 | 218 | 268 | 847 | 142 |
| 0 | TOTAL | 323792 | 170067 | 191622 | 547596 | 807885 | 496126 | 465946 | 490951 | 339609 | 200224 |

Table 3.9 (continued).

|  | Table 1 | Catch numbers at age |  | Numbers*10**-3 |  |  |  | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 3911 | 3407 | 8948 | 3108 | 6942 | 24634 | 28968 | 13648 | 9828 | 5085 |
|  | 4 | 17086 | 9466 | 20933 | 19594 | 14240 | 45769 | 70993 | 137106 | 22774 | 17313 |
|  | 5 | 81986 | 20803 | 19345 | 20473 | 18807 | 27806 | 78672 | 98210 | 135347 | 32165 |
|  | 6 | 40061 | 63433 | 28084 | 17656 | 20086 | 19418 | 25215 | 61407 | 54379 | 81756 |
|  | 7 | 17664 | 21788 | 42496 | 17004 | 15145 | 11369 | 11711 | 13707 | 21015 | 27854 |
|  | 8 | 7442 | 9933 | 8395 | 18329 | 8287 | 3747 | 4063 | 3866 | 3304 | 5501 |
|  | 9 | 3508 | 4267 | 2878 | 2545 | 5988 | 1557 | 976 | 910 | 1236 | 827 |
|  | 10 | 3196 | 1311 | 708 | 646 | 783 | 768 | 726 | 455 | 519 | 290 |
|  | 11 | 678 | 882 | 271 | 229 | 232 | 137 | 557 | 187 | 106 | 41 |
|  | 12 | 79 | 109 | 260 | 74 | 153 | 36 | 136 | 227 | 69 | 13 |
|  | +gp | 58 | 41 | 37 | 83 | 69 | 71 | 76 | 100 | 62 | 28 |
| 0 | TOTALNUM | 175669 | 135440 | 132355 | 99741 | 90732 | 135312 | 222093 | 329823 | 248639 | 170873 |
|  | TONSLAND | 380434 | 399038 | 363730 | 289992 | 277651 | 307920 | 430113 | 523071 | 434939 | 332481 |
|  | SOPCOF \% | 127 | 118 | 125 | 90 | 95 | 102 | 102 | 102 | 100 | 99 |
|  | Table 1 | Catch numbers at age |  | Numbers*10**-3 |  |  |  |  |  |  |  |
|  | YEAR | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 1911 | 4963 | 21835 | 10094 | 6531 | 4879 | 7655 | 12827 | 31887 | 7501 |
|  | 4 | 7551 | 10933 | 36015 | 46182 | 59444 | 42587 | 28782 | 36491 | 88874 | 77714 |
|  | 5 | 12999 | 16467 | 27494 | 63578 | 102548 | 115329 | 80711 | 69633 | 48972 | 92816 |
|  | 6 | 17827 | 20342 | 23392 | 33623 | 59766 | 98485 | 100509 | 83017 | 40493 | 31139 |
|  | 7 | 30007 | 19479 | 18351 | 14866 | 32504 | 32036 | 54590 | 65768 | 34513 | 15778 |
|  | 8 | 6810 | 25193 | 13541 | 9449 | 10019 | 7334 | 10545 | 28392 | 26354 | 15851 |
|  | 9 | 828 | 3888 | 18321 | 6571 | 6163 | 3014 | 2023 | 4651 | 6583 | 8828 |
|  | 10 | 179 | 428 | 2529 | 12593 | 3671 | 1725 | 930 | 1151 | 965 | 1837 |
|  | 11 | 59 | 48 | 264 | 1749 | 7528 | 1174 | 462 | 373 | 197 | 195 |
|  | 12 | 15 | 12 | 82 | 377 | 995 | 1920 | 230 | 213 | 69 | 40 |
|  | +gp | 13 | 4 | 13 | 86 | 144 | 264 | 894 | 383 | 117 | 72 |
| 0 | TOTALNUM | 78199 | 101757 | 161837 | 199168 | 289313 | 308747 | 287331 | 302899 | 279024 | 251771 |
|  | TONSLAND | 212000 | 319158 | 513234 | 581611 | 771086 | 739999 | 732228 | 762403 | 592624 | 484910 |
|  | SOPCOF \% | 101 | 95 | 103 | 101 | 101 | 100 | 101 | 100 | 101 | 100 |
|  | Table 1 | Catch numbers at age |  | Numbers*10**-3 |  |  |  |  |  |  |  |
|  | YEAR | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 4701 | 5044 | 2348 | 7263 | 2090 | 5815 | 8548 | 25473 | 8459 | 4866 |
|  | 4 | 33094 | 35019 | 31033 | 20885 | 38226 | 19768 | 47207 | 43817 | 51704 | 38711 |
|  | 5 | 93044 | 62139 | 76175 | 64447 | 50826 | 113144 | 33625 | 62877 | 40656 | 83998 |
|  | 6 | 47210 | 62456 | 67656 | 71109 | 68350 | 61665 | 78150 | 26303 | 35072 | 46639 |
|  | 7 | 12671 | 22794 | 42122 | 36706 | 50838 | 44777 | 31770 | 34392 | 14037 | 20789 |
|  | 8 | 6677 | 5266 | 11527 | 14002 | 18118 | 20553 | 15667 | 11240 | 20676 | 8417 |
|  | 9 | 4787 | 1773 | 1801 | 2887 | 6239 | 6285 | 7245 | 4080 | 5503 | 8920 |
|  | 10 | 1647 | 1163 | 529 | 492 | 1746 | 2348 | 1788 | 1381 | 1794 | 1957 |
|  | 11 | 321 | 343 | 223 | 142 | 295 | 562 | 737 | 505 | 715 | 872 |
|  | 12 | 71 | 85 | 120 | 97 | 127 | 100 | 210 | 285 | 229 | 987 |
|  | +gp | 26 | 35 | 36 | 65 | 63 | 52 | 226 | 92 | 81 | 117 |
| 0 | TOTALNUM | 204249 | 196117 | 233570 | 218095 | 236918 | 275069 | 225173 | 210445 | 178926 | 216273 |
|  | TONSLAND | 414868 | 426471 | 535045 | 551990 | 606445 | 641276 | 537642 | 486883 | 464171 | 523430 |
|  | SOPCOF \% | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |

Table 3.10. North-East Arctic COD. Catch weights at age

Run title: Arctic Cod (run: SVPASA15/V15) At 22/04/2010 19:53


Table 3.10 (continued).

|  | Table 2 <br> YEAF | Catch weights at age (kg) |  |  | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1980 | 1981 | 1982 |  |  |  |  |  |  |  |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0.27 | 0.49 | 0.37 | 0.84 | 1.42 | 0.94 | 0.64 | 0.49 | 0.54 | 0.74 |
|  | 4 | 0.56 | 0.98 | 0.66 | 1.37 | 1.93 | 1.37 | 1.27 | 0.88 | 0.85 | 0.96 |
|  | 5 | 1.02 | 1.44 | 1.35 | 2.09 | 2.49 | 2.02 | 1.88 | 1.55 | 1.32 | 1.31 |
|  | 6 | 1.72 | 2.09 | 1.99 | 2.86 | 3.14 | 3.22 | 2.79 | 2.33 | 2.24 | 1.92 |
|  | 7 | 3.02 | 2.98 | 2.93 | 3.99 | 3.91 | 4.63 | 4.49 | 3.44 | 3.52 | 2.93 |
|  | 8 | 4.2 | 4.85 | 4.24 | 5.58 | 4.91 | 6.04 | 5.84 | 5.92 | 5.35 | 4.64 |
|  | 9 | 5.84 | 6.57 | 6.46 | 7.77 | 6.02 | 7.66 | 6.83 | 8.6 | 8.06 | 7.52 |
|  | 10 | 7.26 | 9.16 | 8.51 | 9.29 | 7.4 | 9.81 | 7.69 | 9.6 | 9.51 | 9.12 |
|  | 11 | 8.84 | 10.82 | 12.24 | 11.55 | 8.13 | 11.8 | 9.81 | 12.17 | 11.36 | 11.08 |
|  | 12 | 9.28 | 10.77 | 10.78 | 16.2 | 8.57 | 14.16 | 10.71 | 13.72 | 14.09 | 11.47 |
|  | +gp | 14.448 | 13.932 | 14.041 | 17.034 | 8.609 | 14.008 | 12.051 | 13.38 | 16.706 | 16.484 |
| 0 | SOPC | 1.2723 | 1.1809 | 1.2521 | 0.8953 | 0.9483 | 1.0182 | 1.016 | 1.0224 | 1.0001 | 0.9879 |
|  | Table 2 | Catch weights at age (kg) |  |  |  |  |  |  |  |  |  |
|  | YEAF | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0.81 | 1.05 | 1.16 | 0.81 | 0.82 | 0.77 | 0.79 | 0.67 | 0.68 | 0.63 |
|  | 4 | 1.22 | 1.45 | 1.57 | 1.52 | 1.3 | 1.2 | 1.11 | 1.04 | 1.05 | 1.01 |
|  | 5 | 1.64 | 2.15 | 2.21 | 2.16 | 2.06 | 1.78 | 1.61 | 1.53 | 1.62 | 1.54 |
|  | 6 | 2.22 | 2.89 | 3.1 | 2.79 | 2.89 | 2.59 | 2.46 | 2.22 | 2.3 | 2.34 |
|  | 7 | 3.24 | 3.75 | 4.27 | 4.07 | 3.21 | 3.81 | 3.82 | 3.42 | 3.3 | 3.21 |
|  | 8 | 4.68 | 4.71 | 5.19 | 5.53 | 5.2 | 4.99 | 5.72 | 5.2 | 4.86 | 4.29 |
|  | 9 | 7.3 | 6.08 | 6.14 | 6.47 | 6.8 | 6.23 | 6.74 | 7.19 | 6.87 | 6 |
|  | 10 | 9.84 | 8.82 | 7.77 | 7.19 | 7.57 | 8.05 | 8.04 | 7.73 | 9.3 | 6.73 |
|  | 11 | 13.25 | 11.8 | 10.12 | 7.98 | 8.01 | 8.74 | 9.28 | 8.61 | 10.3 | 10.08 |
|  | 12 | 16.88 | 16.58 | 11.54 | 10.11 | 9.48 | 9.22 | 10.4 | 11.07 | 15.05 | 13.88 |
|  | +gp | 11.617 | 16.69 | 14.332 | 14.183 | 11.978 | 12.319 | 10.966 | 11.117 | 14.524 | 14.036 |
| 0 | SOPC | 1.0108 | 0.9521 | 1.027 | 1.0127 | 1.009 | 1.003 | 1.0147 | 1.0004 | 1.0072 | 0.9967 |
|  | Table 2 | Catch weights at age (kg) |  |  |  |  |  |  |  |  |  |
|  | YEAF | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0.572 | 0.66 | 0.723 | 0.672 | 0.72 | 0.693 | 0.721 | 0.736 | 0.769 | 0.747 |
|  | 4 | 1.036 | 1.05 | 1.133 | 1.119 | 1.13 | 1.081 | 1.145 | 1.214 | 1.273 | 1.173 |
|  | 5 | 1.609 | 1.62 | 1.56 | 1.827 | 1.607 | 1.566 | 1.603 | 1.832 | 1.866 | 1.735 |
|  | 6 | 2.344 | 2.51 | 2.306 | 2.499 | 2.429 | 2.205 | 2.388 | 2.511 | 2.818 | 2.419 |
|  | 7 | 3.341 | 3.51 | 3.52 | 3.575 | 3.274 | 3.263 | 3.318 | 3.822 | 3.786 | 3.864 |
|  | 8 | 4.476 | 4.78 | 4.784 | 5.039 | 4.725 | 4.443 | 4.535 | 5.043 | 5.122 | 5.346 |
|  | 9 | 5.724 | 6.04 | 6.2 | 6.355 | 6.712 | 6.228 | 5.466 | 6.584 | 6.223 | 6.428 |
|  | 10 | 7.523 | 7.54 | 7.659 | 8.196 | 7.984 | 8.187 | 6.777 | 8.077 | 7.752 | 8.008 |
|  | 11 | 8.021 | 9 | 9.14 | 10.711 | 9.192 | 9.724 | 7.699 | 8.942 | 8.405 | 8.667 |
|  | 12 | 12.478 | 10.48 | 8.197 | 11.958 | 12.024 | 11.496 | 8.578 | 10.173 | 10.117 | 8.547 |
|  | +gp | 17.241 | 16.18 | 10.325 | 10.657 | 14.245 | 14.417 | 10.155 | 13.364 | 13.674 | 12.022 |
| 0 | SOPC | 1.0039 | 0.9994 | 1.0025 | 1.0014 | 1.0017 | 0.9993 | 0.9981 | 0.9978 | 1.0011 | 1.0002 |

Table 3.11. North-East Arctic COD. Stock weights at age


Table 3.11 (continued).


Table 3.11 (continued).

| Table 3 Stock weights at age (kg) |  |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| YEAF | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 3 | 0.194 | 0.285 | 0.251 | 0.23 | 0.25 | 0.231 | 0.256 | 0.262 | 0.286 | 0.26 |
| 4 | 0.465 | 0.522 | 0.605 | 0.537 | 0.546 | 0.624 | 0.602 | 0.699 | 0.734 | 0.641 |
| 5 | 1.208 | 1.196 | 1.189 | 1.31 | 1.087 | 1.118 | 1.201 | 1.341 | 1.37 | 1.343 |
| 6 | 1.972 | 2.239 | 2.138 | 2.009 | 2.035 | 1.932 | 2.009 | 2.121 | 2.367 | 2.36 |
| 7 | 3.048 | 3.313 | 3.333 | 3.241 | 2.921 | 3.046 | 3.114 | 3.167 | 3.29 | 3.763 |
| 8 | 4.096 | 5.118 | 4.766 | 4.971 | 4.384 | 3.955 | 4.427 | 4.64 | 4.82 | 5.111 |
| 9 | 5.724 | 6.376 | 6.859 | 6.739 | 6.254 | 5.811 | 6.03 | 6.495 | 6.548 | 6.554 |
| 10 | 7.457 | 9.241 | 9.333 | 8.706 | 8.543 | 8.289 | 8.037 | 9.123 | 8.483 | 9.098 |
| 11 | 9.582 | 11.322 | 10.186 | 15.026 | 9.735 | 13.44 | 9.928 | 11.78 | 8.902 | 9.432 |
| 12 | 12.731 | 12.731 | 12.731 | 12.731 | 12.731 | 12.731 | 12.731 | 12.731 | 12.731 | 12.731 |
| +gp | 14.311 | 14.311 | 14.311 | 14.311 | 14.311 | 14.311 | 14.311 | 14.311 | 14.311 | 14.311 |
| 1 |  |  |  |  |  |  |  |  |  |  |

Table 3.12. Northeast Arctic cod. Proportion mature at age.


Table 3.12 (continued)


Table 3.12 (continued)

| Table 5 Proportion mature at age |  |  |  |  |  | $2005$ | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAF | 2000 | 2001 | 2002 | 2003 | 2004 |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 3 | 0 | 0.003 | 0.002 | 0.001 | 0.001 | 0 | 0 | 0 | 0 | 0 |
| 4 | 0.001 | 0.003 | 0.013 | 0.001 | 0.01 | 0.004 | 0.001 | 0.004 | 0.004 | 0 |
| 5 | 0.071 | 0.065 | 0.084 | 0.088 | 0.091 | 0.068 | 0.06 | 0.072 | 0.062 | 0.076 |
| 6 | 0.247 | 0.359 | 0.388 | 0.326 | 0.442 | 0.397 | 0.369 | 0.343 | 0.282 | 0.372 |
| 7 | 0.643 | 0.624 | 0.683 | 0.672 | 0.726 | 0.716 | 0.647 | 0.723 | 0.538 | 0.755 |
| 8 | 0.83 | 0.819 | 0.841 | 0.888 | 0.872 | 0.892 | 0.897 | 0.876 | 0.863 | 0.857 |
| 9 | 0.978 | 0.952 | 0.951 | 0.957 | 0.976 | 0.967 | 0.965 | 0.976 | 0.928 | 0.977 |
| 10 | 1 | 1 | 1 | 1 | 0.977 | 0.991 | 1 | 1 | 0.994 | 0.997 |
| 11 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 12 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| +gp | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1 |  |  |  |  |  |  |  |  |  |  |

Table 3.13. North-East Arctic COD. Tuning data

| North-East Arctic |  | cod | (Sub-areas I |  | and | II) | (run | name: | XSAASA01) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| FLT09: | Russian |  | trawl | catch | and | effort | ages | 9 | - | 11 | (Catch: | Thousa | (Catch: | Unknown) | (Effort: | Unknown) |
| 1985 | 2009 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 9 | 11 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0.7 | 291 | 77 | 30 |  |  |  |  |  |  |  |  |  |  |  |
|  | 1.52 | 87 | 59 | 22 |  |  |  |  |  |  |  |  |  |  |  |
|  | 2.1 | 127 | 95 | 37 |  |  |  |  |  |  |  |  |  |  |  |
|  | 2.75 | 442 | 215 | 53 |  |  |  |  |  |  |  |  |  |  |  |
|  | 2.12 | 140 | 47 | 11 |  |  |  |  |  |  |  |  |  |  |  |
|  | 1.11 | 204 | 49 | 14 |  |  |  |  |  |  |  |  |  |  |  |
|  | 1.56 | 791 | 71 | 16 |  |  |  |  |  |  |  |  |  |  |  |
|  | 2.5 | 3852 | 689 | 62 |  |  |  |  |  |  |  |  |  |  |  |
|  | 2.64 | 2019 | 1778 | 68 |  |  |  |  |  |  |  |  |  |  |  |
|  | 2.96 | 1237 | 595 | 167 |  |  |  |  |  |  |  |  |  |  |  |
|  | 3.88 | 684 | 345 | 146 |  |  |  |  |  |  |  |  |  |  |  |
|  | 3.73 | 364 | 164 | 34 |  |  |  |  |  |  |  |  |  |  |  |
|  | 4.92 | 488 | 99 | 34 |  |  |  |  |  |  |  |  |  |  |  |
|  | 6.77 | 559 | 88 | 34 |  |  |  |  |  |  |  |  |  |  |  |
|  | 6.39 | 882 | 171 | 0 |  |  |  |  |  |  |  |  |  |  |  |
|  | 4.25 | 742 | 185 | 25 |  |  |  |  |  |  |  |  |  |  |  |
|  | 3.5 | 235 | 95 | 35 |  |  |  |  |  |  |  |  |  |  |  |
|  | 3.15 | 336 | 61 | 18 |  |  |  |  |  |  |  |  |  |  |  |
|  | 2.34 | 319 | 83 | 19 |  |  |  |  |  |  |  |  |  |  |  |
|  | 3.47 | 710 | 262 | 56 |  |  |  |  |  |  |  |  |  |  |  |
|  | 3.54 | 588 | 203 | 57 |  |  |  |  |  |  |  |  |  |  |  |
|  | 3.64 | 1182 | 183 | 102 |  |  |  |  |  |  |  |  |  |  |  |
|  | 2.69 | 554 | 244 | 83 |  |  |  |  |  |  |  |  |  |  |  |
|  | 2 | 1741 | 556 | 175 |  |  |  |  |  |  |  |  |  |  |  |
|  | 2.05 | 909 | 520 | 133 |  |  |  |  |  |  |  |  |  |  |  |
| FLT15: | NorBarTrS | rev99 | (Catch: | Unknown) | (Effort: | Unknown) |  |  |  |  |  |  |  |  |  |
| 1980 | 2009 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.99 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 3 | 8 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 233 | 400 | 384 | 48 | 10 | 3 |  |  |  |  |  |  |  |  |
|  | 1 | 277 | 236 | 155 | 160 | 14 | 2 |  |  |  |  |  |  |  |  |
|  | 1 | 523 | 433 | 170 | 58 | 32 | 10 |  |  |  |  |  |  |  |  |
|  | 1 | 283 | 214 | 117 | 41 | 4 | 1 |  |  |  |  |  |  |  |  |
|  | 1 | 1260 | 199 | 77 | 33 | 2 | 1 |  |  |  |  |  |  |  |  |
|  | 1 | 1439 | 641 | 83 | 19 | 3 | 0 |  |  |  |  |  |  |  |  |
|  | 1 | 3911 | 543 | 157 | 20 | 5 | 0 |  |  |  |  |  |  |  |  |
|  | 1 | 805 | 1733 | 205 | 36 | 5 | 0 |  |  |  |  |  |  |  |  |
|  | 1 | 759 | 378 | 902 | 98 | 9 | 1 |  |  |  |  |  |  |  |  |
|  | 1 | 349 | 346 | 206 | 272 | 16 | 4 |  |  |  |  |  |  |  |  |
|  | 1 | 337 | 257 | 215 | 122 | 127 | 6 |  |  |  |  |  |  |  |  |
|  | 1 | 577 | 178 | 128 | 77 | 43 | 27 |  |  |  |  |  |  |  |  |
|  | 1 | 1401 | 725 | 158 | 62 | 39 | 22 |  |  |  |  |  |  |  |  |
|  | 1 | 3102 | 1474 | 506 | 93 | 24 | 16 |  |  |  |  |  |  |  |  |
|  | 1 | 2414 | 2559 | 767 | 185 | 24 | 8 |  |  |  |  |  |  |  |  |
|  | 1 | 1154 | 1372 | 1061 | 240 | 29 | 4 |  |  |  |  |  |  |  |  |
|  | 1 | 640 | 704 | 527 | 283 | 57 | 9 |  |  |  |  |  |  |  |  |
|  | 1 | 1813 | 365 | 259 | 178 | 86 | 10 |  |  |  |  |  |  |  |  |
|  | 1 | 1732 | 581 | 134 | 65 | 51 | 12 |  |  |  |  |  |  |  |  |
|  | 1 | 1321 | 1083 | 269 | 43 | 20 | 12 |  |  |  |  |  |  |  |  |
|  | 1 | 1828 | 834 | 382 | 89 | 11 | 4 |  |  |  |  |  |  |  |  |
|  | 1 | 1350 | 1096 | 425 | 151 | 24 | 3 |  |  |  |  |  |  |  |  |
|  | 1 | 1297 | 911 | 673 | 183 | 49 | 10 |  |  |  |  |  |  |  |  |
|  | 1 | 1725 | 569 | 447 | 273 | 76 | 17 |  |  |  |  |  |  |  |  |
|  | 1 | 621 | 981 | 247 | 155 | 45 | 11 |  |  |  |  |  |  |  |  |
|  | 1 | 1115 | 287 | 437 | 102 | 49 | 14 |  |  |  |  |  |  |  |  |
|  | 1 | 850 | 629 | 148 | 179 | 48 | 18 |  |  |  |  |  |  |  |  |
|  | 1 | 3336 | 910 | 472 | 130 | 88 | 20 |  |  |  |  |  |  |  |  |
|  | 1 | 2196 | 1939 | 586 | 196 | 68 | 49 |  |  |  |  |  |  |  |  |
|  | 1 | 1069 | 1608 | 1407 | 400 | 119 | 35 |  |  |  |  |  |  |  |  |

Table 3.13 (continued)

| FLT16: | NorBarLof/ | rev99 | (Catch: | Unknown) | (Effort: | Unknown) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 2009 |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.99 | 1 |  |  |  |  |  |  |  |  |
| 3 | 9 |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 1416 | 204 | 154 | 157 | 33 | 13 | 10 |  |  |  |
|  | 1 | 1343 | 684 | 116 | 77 | 31 | 3 | 0 |  |  |  |
|  | 1 | 2049 | 502 | 174 | 14 | 30 | 7 | 0 |  |  |  |
|  | 1 | 355 | 578 | 109 | 40 | 3 | 0 | - 1 |  |  |  |
|  | 1 | 344 | 214 | 670 | 166 | 32 | 5 | 2 |  |  |  |
|  | 1 | 206 | 262 | 269 | 668 | 73 | 6 | 3 |  |  |  |
|  | 1 | 346 | 293 | 339 | 367 | 500 | 37 | , |  |  |  |
|  | 1 | 658 | 215 | 184 | 284 | 254 | 824 | 43 |  |  |  |
|  | 1 | 1911 | 1131 | 354 | 255 | 252 | 277 | 442 |  |  |  |
|  | 1 | 4045 | 2175 | 895 | 225 | 119 | 94 | 39 |  |  |  |
|  | 1 | 1598 | 2166 | 1040 | 290 | 44 | 43 | 30 |  |  |  |
|  | 1 | 705 | 872 | 891 | 446 | 65 | 11 | 4 |  |  |  |
|  | 1 | 517 | 497 | 422 | 499 | 205 | 22 | 5 |  |  |  |
|  | 1 | 1826 | 424 | 338 | 340 | 247 | 49 | - 7 |  |  |  |
|  | 1 | 964 | 454 | 122 | 112 | 187 | 92 | 10 |  |  |  |
|  | 1 | 1589 | 1457 | 493 | 129 | 69 | 52 | 12 |  |  |  |
|  | 1 | 1716 | 816 | 573 | 198 | 24 | 8 | 6 |  |  |  |
|  | 1 | 1122 | 1043 | 661 | 345 | 95 | 12 | 5 |  |  |  |
|  | 1 | 1144 | 1315 | 1445 | 643 | 212 | 38 | 5 |  |  |  |
|  | 1 | 928 | 327 | 451 | 468 | 222 | 88 | 22 |  |  |  |
|  | 1 | 337 | 661 | 299 | 432 | 172 | 75 | 18 |  |  |  |
|  | 1 | 591 | 157 | 381 | 169 | 155 | 88 | 24 |  |  |  |
|  | 1 | 371 | 318 | 130 | 426 | 137 | 75 | 35 |  |  |  |
|  | 1 | 3061 | 1410 | 754 | 246 | 329 | 58 | 28 |  |  |  |
|  | 1 | 1783 | 1405 | 495 | 401 | 133 | 260 | 37 |  |  |  |
|  | 1 | 1219 | 1759 | 1949 | 709 | 375 | 111 | 88 |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| FLT18: | RusSwept, | rev05 | (ages | 3-9) | (Catch: | Unknown) | ( | (Catch: | Unknown) | (Effort: | Unknown) |
| 1982 | 2009 |  |  |  |  |  |  |  |  |  |  |
| 1 | 1 | 0.9 | 1 |  |  |  |  |  |  |  |  |
| 3 | 9 |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 1413 | 1525 | 721 | 198 | 551 | 174 | 37 |  |  |  |
|  | 1 | 520 | 642 | 506 | 358 | 179 | 252 | 94 |  |  |  |
|  | 1 | 1189 | 700 | 489 | 357 | 154 | 69 | 61 |  |  |  |
|  | 1 | 1188 | 1592 | 1068 | 365 | 165 | 37 | 8 |  |  |  |
|  | 1 | 1622 | 1532 | 1493 | 481 | 189 | 42 | 2 |  |  |  |
|  | 1 | 557 | 3076 | 900 | 701 | 184 | 60 | 25 |  |  |  |
|  | 1 | 993 | 938 | 2879 | 583 | 260 | 47 | 24 |  |  |  |
|  | 1 | 490 | 978 | 1062 | 1454 | 1167 | 299 | 112 |  |  |  |
|  | 1 | 167 | 487 | 627 | 972 | 1538 | 673 | 153 |  |  |  |
|  | 1 | 1077 | 484 | 532 | 583 | 685 | 747 | 98 |  |  |  |
|  | 1 | 675 | 308 | 239 | 273 | 218 | 175 | 25 |  |  |  |
|  | 1 | 1604 | 1135 | 681 | 416 | 354 | 87 | 3 |  |  |  |
|  | 1 | 1363 | 1309 | 1019 | 354 | 128 | 49 | 21 |  |  |  |
|  | 1 | 589 | 1065 | 1395 | 849 | 251 | 83 | 19 |  |  |  |
|  | 1 | 733 | 784 | 1035 | 773 | 348 | 132 | 19 |  |  |  |
|  | 1 | 1342 | 835 | 613 | 602 | 348 | 116 | 32 |  |  |  |
|  | 1 | 2028 | 1363 | 788 | 470 | 259 | 130 | 48 |  |  |  |
|  | 1 | 1587 | 2072 | 980 | 301 | 123 | 94 | 42 |  |  |  |
|  | 1 | 1839 | 1286 | 1786 | 773 | 114 | 52 | 23 |  |  |  |
|  | 1 | 1224 | 1557 | 1290 | 1061 | 304 | 50 | 14 |  |  |  |
|  | 1 | 980 | 1473 | 1473 | 896 | 600 | 182 | 29 |  |  |  |
|  | 1 | 1246 | 1057 | 1166 | 1203 | 535 | 241 | 40 |  |  |  |
|  | 1 | 329 | 1576 | 880 | 1111 | 776 | 279 | 93 |  |  |  |
|  | 1 | 1408 | 631 | 1832 | 744 | 605 | 244 | 88 |  |  |  |
|  | 1 | 927 | 1613 | 777 | 1801 | 662 | 342 | 161 |  |  |  |
|  | 1 | 2579 | 1617 | 1903 | 846 | 1525 | 553 | 226 |  |  |  |
|  | 1 | 2203 | 3088 | 1635 | 1472 | 830 | 863 | 291 |  |  |  |
|  | 1 | 974 | 2317 | 3687 | 2016 | 1175 | 620 | 413 |  |  |  |

Table 3.14. Northeast arctic cod. Final xsa compared with single fleet tunings run with standard shrinkage settings. Upper part of table shows the weight given to shrinkage at the various runs. Pshrink is population shrinkage and Fshrink is F-shrinkage. Values above 0.3 are shown in bold. Lower part of the table shows population and $F$ at age as estimated before shrinkage (prediction values listed in xsa diagnostics) compared to final run (ALL) with shrinkage. Fs for the youngest ages (3-5) includes cannibalism mortality. *Fleet 007 was not included in the final run.

|  |  | FLT 09 <br> Rus trawl <br> CPUE | FLT 15 <br> Joint BT <br> survey | FLT 16 <br> Joint+Lof <br> Ac survey | FLT 18 <br> Rus BT <br> survey | FLT 007 <br> Joint Eco * <br> BT <br> survey | Final run <br> ALL <br> Fleets |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages with fleet data |  | 9 to 11 | 3 to 8 | 3 to 9 | 3 to 9 | 3 to 11 | 3 to 11 |
| age3 | PshrinkW | 0.91 | 0.52 | 0.51 | 0.54 | 0.48 | 0.22 |
| age4 | FshrinkW | 0.09 | 0.05 | 0.06 | 0.04 | 0.05 | 0.03 |
|  | PshrinkW | 0.89 | 0.38 | 0.35 | 0.40 | 0.33 | 0.14 |
|  | FshrinkW | 0.11 | 0.03 | 0.04 | 0.03 | 0.04 | 0.02 |
| age5 | PshrinkW | 0.87 | 0.39 | 0.39 | 0.43 | 0.33 | 0.16 |
|  | FshrinkW | 0.13 | 0.03 | 0.04 | 0.03 | 0.03 | 0.01 |
| age6 | FshrinkW | 1.00 | 0.04 | 0.04 | 0.04 | 0.04 | 0.01 |
| age7 | FshrinkW | 1.00 | 0.04 | 0.04 | 0.04 | 0.04 | 0.01 |
| age8 | FshrinkW | 1.00 | 0.04 | 0.05 | 0.03 | 0.04 | 0.01 |
| age9 | FshrinkW | 0.14 | 0.10 | 0.05 | 0.04 | 0.06 | 0.02 |
| age10 | FshrinkW | 0.09 | 0.24 | 0.14 | 0.06 | 0.05 | 0.03 |
| age11 | FshrinkW | 0.07 | 0.40 | 0.21 | 0.10 | 0.10 | 0.04 |
| age12 | FshrinkW | 0.10 | 0.53 | 0.27 | 0.17 | 0.17 | 0.08 |
| 2009 | F(5-10) | 0.496 | 0.360 | 0.373 | 0.255 | 0.343 | 0.276 |
| $\begin{aligned} & \text { TSB2009 } \\ & \text { SSB2009 } \end{aligned}$ | incl Age1-2 | 1788 | 2437 | 2383 | 2710 | 2220 | 2782 |
|  | ('000 T) | 725 | 941 | 903 | 1197 | 871 | 1077 |
| N2010 <br> $\mathrm{N}^{*} 10^{\wedge}-3$ <br> with <br> shrinkage | yc2007 | 455027 | 554053 | 550594 | 568507 | 513105 | 592242 |
|  | yc2006 | 328920 | 383236 | 424956 | 402101 | 338144 | 424117 |
|  | yc2005 | 233356 | 340066 | 352407 | 339536 | 314306 | 418565 |
|  | yc2004 | 145304 | 228717 | 219978 | 230583 | 217330 | 298379 |
|  | yc2003 | 63245 | 137799 | 129958 | 133276 | 100273 | 144949 |
|  | yc2002 | 22045 | 62301 | 58441 | 73049 | 48191 | 72212 |
|  | yc2001 | 7814 | 23874 | 15425 | 28350 | 15061 | 24121 |
|  | yc2000 | 9783 | 12465 | 15839 | 22325 | 16199 | 17680 |
|  |  | No | shrinkage |  |  |  | Shrinkage |
| Survivors end of 09 <br> direct <br> predic. <br> by the <br> survey <br> $\mathrm{N}^{*} 10^{\wedge}-3$ | yc2006 |  | 353082 | 459119 | 378731 | 293981 | 424117 |
|  | yc2005 |  | 383509 | 410343 | 376000 | 344108 | 418565 |
|  | yc2004 |  | 278913 | 268260 | 279021 | 255536 | 298379 |
|  | yc2003 |  | 141014 | 133655 | 135596 | 101840 | 144949 |
|  | yc2002 |  | 64832 | 61093 | 75723 | 49934 | 72212 |
|  | yc2001 |  | 25034 | 16057 | 29463 | 15717 | 24121 |
|  | yc2000 | 9868 | 13221 | 16519 | 23076 | 16968 | 17680 |
|  | yc1999 | 13491 | 2443 | 2359 | 9340 | 6211 | 5224 |
| F2009 | yc2006 |  | 0.154 | 0.121 | 0.144 | 0.183 | 0.13 |
|  | yc2005 |  | 0.105 | 0.099 | 0.107 | 0.117 | 0.097 |
| direct <br> predic. <br> by the survey | yc2004 |  | 0.241 | 0.250 | 0.241 | 0.261 | 0.227 |
|  | yc2003 |  | 0.262 | 0.274 | 0.271 | 0.347 | 0.256 |
|  | yc2002 |  | 0.255 | 0.268 | 0.222 | 0.320 | 0.232 |
|  | yc2001 |  | 0.266 | 0.388 | 0.230 | 0.395 | 0.275 |
|  | yc2000 | 0.598 | 0.477 | 0.398 | 0.300 | 0.389 | 0.376 |
|  | yc1999 | 0.123 | 0.545 | 0.560 | 0.174 | 0.251 | 0.292 |
| 2009 | F(5-10) |  | 0.341 | 0.356 | 0.240 | 0.327 | 0.276 |

Table 3.15. Northeast Arctic Cod. Diagnostics for final XSA.

| Lowestoft VPA Version 3.1 |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |
| 24/04/2010 17:49 |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Extended Survivors Analysis |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Arctic Cod (run: XSAASA01/X01) |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| CPUE data from file fleet |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Catch data for 10 years. 2000 to 2009. Ages 1 to 13. |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Fleet | Fil | Last | First | Last | Alpha | Beta |  |  |  |  |
|  | year | year | age | age |  |  |  |  |  |  |
| FLT09: RL | 2000 | 2009 | 9 | 11 | 0 | 1 |  |  |  |  |
| FLT15: Nc | 2000 | 2009 | 3 | 8 | 0.99 | 1 |  |  |  |  |
| FLT16: Nc | 2000 | 2009 | 3 | 9 | 0.99 | 1 |  |  |  |  |
| FLT18. R | 2000 | 2009 | 3 | 9 | 0.9 | 1 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Time series weights |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Tapered time weighting applied |  |  |  |  |  |  |  |  |  |  |
| Power | 3 over | 10 years |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Catchability analysis : |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Catchability dependent on stock size for ages < 6 |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Regression type $=\mathrm{C}$ |  |  |  |  |  |  |  |  |  |  |
| Minimum of 5 points used for regression |  |  |  |  |  |  |  |  |  |  |
| Survivor estimates shrunk to the population mean for ages $<6$ |  |  |  |  |  | 6 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Catchability independent of age for ages > $=10$ |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Terminal population estimation |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Survivor estimates shrunk towards the mean $F$of the final 5 years or the 2 oldest ages. |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| S.E. of the mean to which the estimates are shrunk $=1.000$ |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Minimum standard error for population |  |  |  |  |  |  |  |  |  |  |
| estimates derived from each fleet $=.300$ |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Prior weighting not applied |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Tuning had not converged after |  |  | 30 iteration |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Total absolute residual between iterations |  |  |  |  |  |  |  |  |  |  |
| 29 and $30=.00316$ |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Final year $F$ values |  |  |  |  |  |  |  |  |  |  |
| Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| Iteration 2 | 1.8804 | 0.1892 | 0.13 | 0.0968 | 0.2274 | 0.2557 | 0.2317 | 0.2747 | 0.3766 | 0.2928 |
| Iteration 3 | 1.8803 | 0.1892 | 0.13 | 0.0968 | 0.2274 | 0.2556 | 0.2316 | 0.2745 | 0.3763 | 0.2923 |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Age | 11 | 12 |  |  |  |  |  |  |  |  |
| Iteration 2 | 0.3391 | 0.6368 |  |  |  |  |  |  |  |  |
| Iteration 3 | 0.3385 | 0.6355 |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Regression weights |  |  |  |  |  |  |  |  |  |  |
|  | 0.02 | 0.116 | 0.284 | 0.482 | 0.67 | 0.82 | 0.921 | 0.976 | 0.997 | 1 |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Fishing mortalities |  |  |  |  |  |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  |  |  |  |  |  |  |  |  |  |  |
| 1 | 1.381 | 0.948 | 0.615 | 1.408 | 0.946 | 1.111 | 0.938 | 0.858 | 1.037 | 1.88 |
| 2 | 0.26 | 0.203 | 0.41 | 0.276 | 0.577 | 0.22 | 0.117 | 0.206 | 0.146 | 0.189 |
| 3 | 0.077 | 0.062 | 0.111 | 0.049 | 0.078 | 0.181 | 0.026 | 0.133 | 0.192 | 0.13 |
| 4 | 0.14 | 0.117 | 0.106 | 0.072 | 0.103 | 0.117 | 0.146 | 0.125 | 0.147 | 0.097 |
| 5 | 0.411 | 0.285 | 0.288 | 0.276 | 0.257 | 0.391 | 0.253 | 0.282 | 0.163 | 0.227 |
| 6 | 0.605 | 0.52 | 0.557 | 0.472 | 0.53 | 0.56 | 0.487 | 0.321 | 0.251 | 0.256 |
| 7 | 0.753 | 0.673 | 0.807 | 0.68 | 0.748 | 0.82 | 0.633 | 0.411 | 0.284 | 0.232 |
| 8 | 1.038 | 0.845 | 0.898 | 0.701 | 0.885 | 0.796 | 0.783 | 0.48 | 0.467 | 0.275 |
| 9 | 1.188 | 0.894 | 0.81 | 0.589 | 0.806 | 0.924 | 0.743 | 0.474 | 0.46 | 0.376 |
| 10 | 1.156 | 1.128 | 0.748 | 0.538 | 0.898 | 0.842 | 0.751 | 0.296 | 0.395 | 0.292 |
| 11 | 1.078 | 0.807 | 0.672 | 0.454 | 0.738 | 0.849 | 0.706 | 0.488 | 0.246 | 0.338 |
| 12 | 1.13 | 0.986 | 0.755 | 0.712 | 0.986 | 0.602 | 0.943 | 0.663 | 0.428 | 0.636 |

Table 3.15 (continued)


Table 3.15 (continued)


Table 3.15 (continued)

| Fleet : FLT18: RusSweptArea |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 3 | 0.19 | 0.14 | 0.18 | -0.16 | -0.01 | 0.16 | -0.12 | 0.02 | 0.05 | -0.07 |
| 4 | -0.04 | -0.08 | 0.02 | -0.04 | -0.24 | -0.01 | 0.12 | -0.02 | 0.1 | 0 |
| 5 | 0.15 | -0.06 | -0.14 | -0.18 | -0.23 | -0.09 | 0.09 | 0.22 | -0.12 | 0.17 |
| 6 | 0.14 | -0.03 | -0.18 | -0.14 | -0.03 | -0.27 | 0.21 | 0.05 | 0.03 | 0.08 |
| 7 | -0.57 | -0.33 | -0.01 | -0.22 | -0.05 | -0.04 | 0.04 | 0.25 | 0.11 | -0.16 |
| 8 | -0.48 | -0.59 | 0.01 | -0.25 | -0.04 | -0.45 | 0.14 | 0.31 | 0.11 | 0.05 |
| 9 | -0.76 | -0.71 | -0.14 | -0.72 | -0.22 | -0.08 | 0.07 | 0.4 | 0.31 | -0.06 |
| 10 |  |  |  |  |  |  |  |  |  |  |
| 11 | No data for this fleet at this age No data for this fleet at this age |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Mean $\log$ catchability and standard error of ages with catchability independent of year class strength and constant w.r.t. time |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Age | 6 | 7 | 8 | 9 |  |  |  |  |  |  |
| Mean Log | -4.3816 | -3.9758 | -3.7383 | -3.7231 |  |  |  |  |  |  |
| S.E(Log q | 0.1563 | 0.1671 | 0.2632 | 0.3419 |  |  |  |  |  |  |
| - |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Regression statistics |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Ages with | q dependent | nt on year c | lass streng |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Age | Slope | $t$-value | Intercept | RSquare | No Pts | Regs.e | Mean Log |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| 3 | 0.53 | 2.998 | 9.39 | 0.91 | 10 | 0.12 | -5.92 |  |  |  |
| 4 | 0.71 | 1.901 | 7.55 | 0.91 | 10 | 0.12 | -5.31 |  |  |  |
| 5 | 0.71 | 1.144 | 7.07 | 0.79 | 10 | 0.2 | -4.83 |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Ages with q independent of year |  |  | class stren | gth and con | nstant w.r.t | time. |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Age | Slope | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| 6 | 0.89 | 0.487 | 5.27 | 0.81 | 10 | 0.15 | -4.38 |  |  |  |
| 7 | 0.96 | 0.148 | 4.31 | 0.72 | 10 | 0.18 | -3.98 |  |  |  |
| 8 | 0.88 | 0.36 | 4.52 | 0.69 | 10 | 0.25 | -3.74 |  |  |  |
| 9 | 0.82 | 0.831 | 4.78 | 0.83 | 10 | 0.29 | -3.72 |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Terminal year survivor and F summaries : |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Age 1 Catchability dependent |  |  | on age and | year class | strength |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Year class $=2008$ |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Fleet |  | Int | Ext | Var | N | Scaled | Estimated |  |  |  |
|  |  | s.e | s.e | Ratio |  | Weights |  |  |  |  |
| FLT09: RL | 1 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |
| FLT15: Nc | 1 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |
| FLT16: Nc | 1 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |
| FLT18: Rl | 1 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| P shrink | 972097 | 0.3 |  |  |  | 0.917 | 1.973 |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| F shrink: | 3564466 | 1 |  |  |  | 0.083 | 0.989 |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Weighted prediction |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |  |  |  |
| at end of ) | s.e | s.e |  | Ratio |  |  |  |  |  |  |
| 1083239 | 0.29 | 13.9 | 2 | 48.156 | 1.88 |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| 1 |  |  |  |  |  |  |  |  |  |  |
| Age 2 Catchability |  | dependent | on age and | year class | strength |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Year class $=2007$ |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Fleet |  | Int | Ext | Var | N | Scaled | Estimated |  |  |  |
|  |  | s.e | s.e | Ratio |  | Weights | F |  |  |  |
| FLT09: Ru | 1 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |
| FLT15: Nc | 1 | 0 | , | 0 | 0 | 0 | 0 |  |  |  |
| FLT16: Nc | , | , | 0 | 0 | 0 | 0 | 0 |  |  |  |
| FLT18: Rı | , | 0 | 0 | 0 | 0 | . | 0 |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| P shrink | 615780 | 0.34 |  |  |  | 0.895 | 0.183 |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| F shrink: | 425486 | 1 |  |  |  | 0.105 | 0.254 |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Weighted prediction |  |  |  |  |  |  |  |  |  |  |
| Survivors |  |  |  |  |  |  |  |  |  |  |
| at end of$592242$ | $s . e^{\operatorname{lnt}}$ | Ext | N | Var | F |  |  |  |  |  |
|  | s.e 0.32 | s.e 13.29 | 2 | Ratio 40.936 | 0.189 |  |  |  |  |  |

Table 3.15 (continued)


## Table 3.15 (continued)



## Table 3.15 (continued)



Table 3.16. Northeast Arctic cod. Fishing mortality for XSA run down to age 1. Number of cod eaten by cod included in catch matrix


Table 3.17. Northeast Arctic cod. Stock number at age


Table 3.18. Northeast Arctic cod. Natural mortality used in final VPA.


Table 3.18 (continued).

| Table 4 Natural Mortality (M) at age |  |  |  |  |  | 1985 |  | 1987 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAF <br> AGE | 1980 | 1981 | 1982 | 1983 | 1984 |  | 1986 |  | 1988 | 1989 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2006 | 0.2004 | 0.3123 | 0.2587 | 0.2087 | 0.2 |
| 4 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 5 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 6 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 7 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 8 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 9 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 10 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 11 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 12 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| +gp | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
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|  |  |  |  |  |  |  |  |  |  |  |
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|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Table | 4 | $1991$ | (M) at age | 1993 |  |  |  |  |  |  |
| YEAF | 1990 |  |  |  | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
|  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 3 | 0.2 | 0.2049 | 0.2067 | 0.2664 | 0.4031 | 0.6726 | 0.6477 | 0.5141 | 0.5276 | 0.3095 |
| 4 | 0.2 | 0.2 | 0.2 | 0.2028 | 0.2929 | 0.3613 | 0.4327 | 0.2935 | 0.2768 | 0.2111 |
| 5 | 0.2 | 0.2 | 0.2 | 0.2024 | 0.2259 | 0.2111 | 0.2812 | 0.2103 | 0.2164 | 0.2 |
| 6 | 0.193 | 0.2 | 0.2 | 0.2 | 0.2047 | 0.2014 | 0.206 | 0.202 | 0.2096 | 0.2 |
| 7 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 8 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 9 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 10 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 11 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 12 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| +gp | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
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|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Table | Natural | Mortality | ) at age |  |  |  |  |  |  |  |
| YEAF | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 3 | 0.2686 | 0.2514 | 0.3051 | 0.237 | 0.2702 | 0.3686 | 0.2096 | 0.3002 | 0.3799 | 0.3203 |
| 4 | 0.2417 | 0.2289 | 0.2162 | 0.2 | 0.2218 | 0.216 | 0.2057 | 0.2113 | 0.2526 | 0.2171 |
| 5 | 0.2168 | 0.2079 | 0.2032 | 0.2 | 0.2054 | 0.2179 | 0.2005 | 0.2001 | 0.2147 | 0.2005 |
| 6 | 0.2006 | 0.2072 | 0.2002 | 0.2 | 0.2003 | 0.2048 | 0.2 | 0.2 | 0.2 | 0.2 |
| 7 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 8 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 9 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 10 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 11 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 12 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| +gp | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 1 |  |  |  |  |  |  |  |  |  |  |

Table 3.19 Northeast arctic cod. Natural mortality of cod (M2) due to cannibalism
Year M2 age $1 \quad$ M2 age 2 M2 age $3 \quad$ M2 age $4 \quad$ M2 age $5 \quad$ M2 age 6

|  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0.2457 | 0.0356 | 0.0006 | 0.0000 | 0.0000 | 0.0000 |
| 1985 | 0.3590 | 0.0562 | 0.0004 | 0.0000 | 0.0000 | 0.0000 |
| 1986 | 0.9368 | 0.8010 | 0.1123 | 0.0000 | 0.0000 | 0.0000 |
| 1987 | 0.5266 | 0.8017 | 0.0584 | 0.0000 | 0.0000 | 0.0000 |
| 1988 | 0.8044 | 0.1094 | 0.0087 | 0.0000 | 0.0000 | 0.0000 |
| 1989 | 0.2145 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1990 | 0.0961 | 0.0590 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1991 | 0.1038 | 0.2373 | 0.0050 | 0.0000 | 0.0000 | 0.0000 |
| 1992 | 0.4681 | 0.1450 | 0.0067 | 0.0000 | 0.0000 | 0.0000 |
| 1993 | 2.5644 | 0.4482 | 0.0660 | 0.0028 | 0.0024 | 0.0000 |
| 1994 | 1.7157 | 0.6312 | 0.1997 | 0.0940 | 0.0259 | 0.0047 |
| 1995* | 1.8681 | 0.9350 | 0.5392 | 0.1859 | 0.0111 | 0.0014 |
| 1996 | 1.9892 | 1.0545 | 0.4450 | 0.2321 | 0.0812 | 0.0060 |
| 1997 | 2.5175 | 1.0927 | 0.3145 | 0.0932 | 0.0103 | 0.0020 |
| 1998 | 1.6230 | 0.6371 | 0.3360 | 0.0795 | 0.0168 | 0.0098 |
| 1999 | 1.1053 | 0.3576 | 0.1022 | 0.0106 | 0.0000 | 0.0000 |
| 2000 | 1.3809 | 0.2596 | 0.0686 | 0.0417 | 0.0168 | 0.0006 |
| 2001 | 0.9477 | 0.2026 | 0.0514 | 0.0289 | 0.0079 | 0.0072 |
| 2002 | 0.6152 | 0.4096 | 0.1051 | 0.0162 | 0.0032 | 0.0002 |
| 2003 | 1.4084 | 0.2760 | 0.0370 | 0.0000 | 0.0000 | 0.0000 |
| 2004 | 0.9464 | 0.5766 | 0.0702 | 0.0218 | 0.0054 | 0.0003 |
| 2005 | 1.1146 | 0.2194 | 0.1686 | 0.0160 | 0.0179 | 0.0048 |
| 2006 | 0.9403 | 0.1161 | 0.0097 | 0.0057 | 0.0005 | 0.0000 |
| 2007 | 0.8616 | 0.2037 | 0.1001 | 0.0113 | 0.0001 | 0.0000 |
| 2008 | 1.0389 | 0.1462 | 0.1806 | 0.0526 | 0.0147 | 0.0000 |
| 2009 | 1.8824 | 0.1886 | 0.1209 | 0.0180 | 0.0006 | 0.0000 |

* corrected data on cod consumption

Table 3.20. Northeast Arctic cod. Fishing mortality, final VPA


Table 3.20 (continued).


Table 3.21 Northeast arctic cod. Fishing mortality of age 1-6 cod

| Year | F age 1 | F age 2 | F age 3 | $F$ age 4 | $F$ age 5 | F age 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0.0000 | 0.0017 | 0.0193 | 0.1235 | 0.3075 | 0.6274 |
| 1985 | 0.0001 | 0.0015 | 0.0529 | 0.1701 | 0.3763 | 0.6051 |
| 1986 | 0.0000 | 0.0017 | 0.0328 | 0.2122 | 0.4933 | 0.7052 |
| 1987 | 0.0000 | 0.0011 | 0.0555 | 0.2285 | 0.5097 | 0.9363 |
| 1988 | 0.0000 | 0.0009 | 0.0542 | 0.1275 | 0.3704 | 0.5971 |
| 1989 | 0.0000 | 0.0009 | 0.0327 | 0.1284 | 0.2674 | 0.4016 |
| 1990 | 0.0000 | 0.0004 | 0.0086 | 0.0622 | 0.1342 | 0.2317 |
| 1991 | 0.0000 | 0.0007 | 0.0133 | 0.0624 | 0.1875 | 0.3210 |
| 1992 | 0.0004 | 0.0011 | 0.0338 | 0.1265 | 0.2205 | 0.4428 |
| 1993 | 0.0000 | 0.0006 | 0.0128 | 0.0933 | 0.3441 | 0.4597 |
| 1994 | 0.0000 | 0.0003 | 0.0101 | 0.1058 | 0.3132 | 0.6409 |
| 1995 | 0.0000 | 0.0003 | 0.0109 | 0.1023 | 0.3270 | 0.5757 |
| 1996 | 0.0000 | 0.0006 | 0.0239 | 0.1203 | 0.3305 | 0.5366 |
| 1997 | 0.0000 | 0.0007 | 0.0231 | 0.2061 | 0.5590 | 0.7219 |
| 1998 | 0.0000 | 0.0019 | 0.0496 | 0.2759 | 0.5057 | 0.7704 |
| 1999 | 0.0000 | 0.0004 | 0.0159 | 0.1987 | 0.5481 | 0.7250 |
| 2000 | 0.0000 | 0.0003 | 0.0088 | 0.0980 | 0.3942 | 0.6044 |
| 2001 | 0.0000 | 0.0004 | 0.0110 | 0.0878 | 0.2773 | 0.5127 |
| 2002 | 0.0001 | 0.0001 | 0.0060 | 0.0900 | 0.2847 | 0.5565 |
| 2003 | 0.0000 | 0.0005 | 0.0117 | 0.0720 | 0.2758 | 0.4718 |
| 2004 | 0.0000 | 0.0002 | 0.0079 | 0.0811 | 0.2517 | 0.5301 |
| 2005 | 0.0000 | 0.0006 | 0.0120 | 0.1012 | 0.3732 | 0.5556 |
| 2006 | 0.0000 | 0.0011 | 0.0168 | 0.1404 | 0.2527 | 0.4866 |
| 2007 | 0.0010 | 0.0026 | 0.0333 | 0.1133 | 0.2821 | 0.3212 |
| 2008 | 0.0000 | 0.0009 | 0.0123 | 0.0949 | 0.1482 | 0.2512 |
| 2009 | 0.0000 | 0.0005 | 0.0097 | 0.0797 | 0.2269 | 0.2556 |

Table 3.22. Northeast Arctic cod. Stock number at age. Final VPA


Table 3.22 (continued).


Table 3.23. Northeast Arctic cod. Stock biomass at age. Final VPA


Table 3.23 (continued).


Table 3.24. Northeast Arctic cod. Spawning stock biomass at age


Table 3.24 (continued).

| Table 13 YEAF | Spawning stock biomass at age (spawning time) |  |  |  |  | $\begin{array}{r\|} \hline \text { Tonnes } \\ 1985 \end{array}$ | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1980 | 1981 | 1982 | 1983 | 1984 |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 3 | 0 | 0 | 0 | 617 | 0 | 0 | 0 | 0 | 0 | 413 |
| 4 | 0 | 0 | 3975 | 8555 | 7759 | 2793 | 17879 | 3663 | 1683 | 245 |
| 5 | 0 | 3205 | 10921 | 12763 | 25258 | 13950 | 25890 | 23594 | 18913 | 3776 |
| 6 | 3248 | 25173 | 48986 | 35756 | 40578 | 47110 | 25408 | 40443 | 82953 | 91093 |
| 7 | 15391 | 24723 | 160097 | 95196 | 51176 | 45665 | 43081 | 16506 | 55598 | 87433 |
| 8 | 23415 | 42732 | 50592 | 129590 | 52586 | 31544 | 27356 | 15555 | 15650 | 33233 |
| 9 | 23759 | 33622 | 26992 | 26639 | 55265 | 23068 | 7669 | 5948 | 15268 | 11403 |
| 10 | 31960 | 17804 | 12436 | 8986 | 10636 | 19184 | 7204 | 4670 | 8256 | 5034 |
| 11 | 8362 | 16843 | 5870 | 5224 | 3521 | 6210 | 11412 | 4142 | 1910 | 941 |
| 12 | 989 | 1899 | 5283 | 1645 | 2794 | 1346 | 2965 | 4496 | 1169 | 330 |
| +gp | 1130 | 924 | 979 | 2209 | 1514 | 2984 | 1863 | 2226 | 1181 | 798 |
| 0 TOTSF | 108253 | 166926 | 326133 | 327181 | 251087 | 193856 | 170729 | 121243 | 202582 | 234698 |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| YEAF | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
|  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 3 | 773 | 213 | 317 | 0 | 551 | 0 | 0 | 0 | 183 | 222 |
| 4 | 1254 | 7162 | 4314 | 18602 | 3566 | 752 | 0 | 0 | 668 | 493 |
| 5 | 6821 | 13758 | 39790 | 37771 | 71178 | 31548 | 5994 | 2307 | 4096 | 3948 |
| 6 | 33899 | 59929 | 81186 | 103102 | 111395 | 193184 | 140083 | 46253 | 24284 | 24994 |
| 7 | 190031 | 139539 | 149505 | 106228 | 115157 | 127794 | 250181 | 256533 | 96921 | 50993 |
| 8 | 60804 | 368315 | 159775 | 116956 | 80973 | 50277 | 89354 | 189281 | 164243 | 99454 |
| 9 | 14670 | 88337 | 359921 | 96591 | 65512 | 36391 | 31427 | 57600 | 73416 | 79512 |
| 10 | 5395 | 22227 | 71933 | 240574 | 46501 | 26131 | 16807 | 19359 | 14955 | 26970 |
| 11 | 1389 | 3966 | 16313 | 41966 | 96272 | 17977 | 9217 | 5908 | 3393 | 3816 |
| 12 | 593 | 944 | 3826 | 10659 | 20437 | 38740 | 5319 | 3767 | 1304 | 795 |
| +gp | 578 | 354 | 682 | 2733 | 3325 | 5988 | 23239 | 7614 | 2485 | 1609 |
| 0 TOTSF | 316206 | 704745 | 887563 | 775183 | 614866 | 528781 | 571620 | 588621 | 385946 | 292807 |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 3 0 442 226 160 75 0 0 0 0 |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| 5 | 26972 | 21879 | 33628 | 33696 | 24799 | 30510 | 11922 | 27133 | 27745 | 46434 |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| 8 | 38183 | 42132 | 84826 | 133631 | 128310 | 144010 | 125189 | 131332 | 252614 | 168683 |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| 11 | 5057 | 7639 | 5063 | 6372 | 6002 | 14351 | 15758 | 16880 | 31906 | 31424 |
| 12 | 1446 | 1874 | 3140 | 2644 | 2800 | 3074 | 4756 | 8166 | 9175 | 29170 |
| +gp | 595 | 867 | 1059 | 1992 | 1561 | 1797 | 5754 | 2963 | 3648 | 3887 |
| 0 TOTSF | 240096 | 354492 | 496423 | 547175 | 654572 | 606902 | 595285 | 649186 | 703780 | 1069646 |
| 1 |  |  |  |  |  |  |  |  |  |  |

Table 3.25. Northeast Arctic cod. Summary Table. Final VPA.

| Run title : Arctic Cod (run: SVPASA15/V15) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| At 22/04/2010 19:53 |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| Table 16 Summary (without SOP correction) |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| Traditional vpa using file input for terminal F |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
|  | RECRUITS | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | FBAR 5-10 |
|  | Age 3 |  |  |  |  |  |
| 1946 | 728139 | 4168882 | 1112776 | 706000 | 0.6344 | 0.1857 |
| 1947 | 425311 | 3692801 | 1165059 | 882017 | 0.7571 | 0.3047 |
| 1948 | 442592 | 3665819 | 1019114 | 774295 | 0.7598 | 0.3398 |
| 1949 | 468348 | 3065111 | 729879 | 800122 | 1.0962 | 0.3619 |
| 1950 | 704908 | 2830103 | 615339 | 731982 | 1.1896 | 0.3566 |
| 1951 | 1083753 | 3141009 | 568705 | 827180 | 1.4545 | 0.3966 |
| 1952 | 1193111 | 3407679 | 520599 | 876795 | 1.6842 | 0.5348 |
| 1953 | 1590377 | 3557376 | 396417 | 695546 | 1.7546 | 0.3572 |
| 1954 | 641584 | 4039204 | 429694 | 826021 | 1.9223 | 0.3879 |
| 1955 | 272778 | 3488383 | 346919 | 1147841 | 3.3087 | 0.5437 |
| 1956 | 439602 | 3189831 | 299823 | 1343068 | 4.4795 | 0.6401 |
| 1957 | 804781 | 2495895 | 207840 | 792557 | 3.8133 | 0.5089 |
| 1958 | 496824 | 2164149 | 195377 | 769313 | 3.9376 | 0.5169 |
| 1959 | 683690 | 2415826 | 432489 | 744607 | 1.7217 | 0.5596 |
| 1960 | 789653 | 2050805 | 383479 | 622042 | 1.6221 | 0.4789 |
| 1961 | 916842 | 2137149 | 404228 | 783221 | 1.9376 | 0.6348 |
| 1962 | 728338 | 1957006 | 311678 | 909266 | 2.9173 | 0.7576 |
| 1963 | 472064 | 1747579 | 208207 | 776337 | 3.7287 | 0.9866 |
| 1964 | 338678 | 1374529 | 186570 | 437695 | 2.346 | 0.6789 |
| 1965 | 776941 | 1440693 | 102315 | 444930 | 4.3486 | 0.5533 |
| 1966 | 1582560 | 2198418 | 120722 | 483711 | 4.0068 | 0.5302 |
| 1967 | 1295416 | 2852164 | 129784 | 572605 | 4.412 | 0.5439 |
| 1968 | 164955 | 3387455 | 227215 | 1074084 | 4.7272 | 0.5704 |
| 1969 | 112039 | 2805591 | 151870 | 1197226 | 7.8832 | 0.8292 |
| 1970 | 197105 | 2057698 | 224482 | 933246 | 4.1573 | 0.7493 |
| 1971 | 404774 | 1610969 | 311662 | 689048 | 2.2109 | 0.5956 |
| 1972 | 1015319 | 1621485 | 346511 | 565254 | 1.6313 | 0.6928 |
| 1973 | 1818949 | 2401955 | 332913 | 792685 | 2.3811 | 0.602 |
| 1974 | 523916 | 2236387 | 164491 | 1102433 | 6.7021 | 0.5633 |
| 1975 | 621616 | 2037430 | 142028 | 829377 | 5.8395 | 0.6595 |
| 1976 | 613942 | 1931396 | 171238 | 867463 | 5.0658 | 0.6457 |
| 1977 | 348054 | 1950748 | 341385 | 905301 | 2.6518 | 0.8379 |
| 1978 | 638490 | 1576565 | 241536 | 698715 | 2.8928 | 0.9406 |
| 1979 | 198490 | 1114381 | 174699 | 440538 | 2.5217 | 0.7264 |
| 1980 | 137735 | 863862 | 108253 | 380434 | 3.5143 | 0.7241 |
| 1981 | 150868 | 983658 | 166926 | 399038 | 2.3905 | 0.8632 |
| 1982 | 151830 | 750871 | 326133 | 363730 | 1.1153 | 0.7583 |
| 1983 | 166831 | 738675 | 327181 | 289992 | 0.8863 | 0.756 |
| 1984 | 397831 | 817596 | 251087 | 277651 | 1.1058 | 0.9161 |
| 1985 | 523674 | 957513 | 193856 | 307920 | 1.5884 | 0.7038 |
| 1986 | 1038825 | 1294449 | 170729 | 430113 | 2.5193 | 0.8649 |
| 1987 | 285293 | 1126053 | 121243 | 523071 | 4.3142 | 0.951 |
| 1988 | 204644 | 915105 | 202582 | 434939 | 2.147 | 0.9743 |
| 1989 | 172782 | 889738 | 234698 | 332481 | 1.4166 | 0.6604 |
| 1990 | 242750 | 961667 | 316206 | 212000 | 0.6704 | 0.2711 |
| 1991 | 411766 | 1561708 | 704745 | 319158 | 0.4529 | 0.321 |
| 1992 | 721185 | 1912257 | 887563 | 513234 | 0.5783 | 0.455 |
| 1993 | 894434 | 2359644 | 775183 | 581611 | 0.7503 | 0.5528 |
| 1994 | 781468 | 2148154 | 614866 | 771086 | 1.2541 | 0.8678 |
| 1995 | 613875 | 1807057 | 528781 | 739999 | 1.3994 | 0.788 |
| 1996 | 438206 | 1687088 | 571620 | 732228 | 1.281 | 0.6987 |
| 1997 | 715163 | 1531894 | 588621 | 762403 | 1.2952 | 1.0338 |
| 1998 | 844814 | 1229614 | 385946 | 592624 | 1.5355 | 0.917 |
| 1999 | 547772 | 1100226 | 292807 | 484910 | 1.6561 | 0.9892 |
| 2000 | 610255 | 1100947 | 240096 | 414868 | 1.7279 | 0.8546 |
| 2001 | 516555 | 1374769 | 354492 | 426471 | 1.203 | 0.7213 |
| 2002 | 449404 | 1544017 | 496423 | 535045 | 1.0778 | 0.684 |
| 2003 | 697062 | 1617723 | 547175 | 551990 | 1.0088 | 0.5441 |
| 2004 | 300161 | 1582559 | 654572 | 606445 | 0.9265 | 0.6866 |
| 2005 | 581337 | 1580766 | 606902 | 641276 | 1.0566 | 0.7172 |
| 2006 | 566885 | 1546829 | 595285 | 537642 | 0.9032 | 0.6063 |
| 2007 | 899016 | 1814225 | 649186 | 486883 | 0.75 | 0.3774 |
| 2008 | 830932 | 2269997 | 703780 | 464171 | 0.6595 | 0.3337 |
| 2009 | 588966 | 2618810 | 1069646 | 523430 | 0.4893 | 0.2762 |
|  |  |  |  |  |  |  |
| Arith. |  |  |  |  |  |  |
| Mean | 609629 | 2007811 | 412557 | 651240 | 2.2527 | 0.6319 |
| 0 Units 1 | (Thousands) | (Tonnes) | (Tonnes) | (Tonnes) |  |  |
|  |  |  |  |  |  |  |

Table 3.26. Northeast Arctic cod. Summary table, run without cannibalism.

| Run title: Arctic Cod (run: SVPASA15/V15) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |
| At 22/04/2010 19:56 |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| Table 16 Summary (without SOP correction) |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| Traditional vpa using file input for terminal F |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
|  | RECRUITS | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | FBAR 5-10 |
|  | Age 3 |  |  |  |  |  |
| 1946 | 728139 | 4168882 | 1112776 | 706000 | 0.6344 | 0.1857 |
| 1947 | 425311 | 3692801 | 1165059 | 882017 | 0.7571 | 0.3047 |
| 1948 | 442592 | 3665819 | 1019114 | 774295 | 0.7598 | 0.3398 |
| 1949 | 468348 | 3065111 | 729879 | 800122 | 1.0962 | 0.3619 |
| 1950 | 704908 | 2830103 | 615339 | 731982 | 1.1896 | 0.3566 |
| 1951 | 1083753 | 3141009 | 568705 | 827180 | 1.4545 | 0.3966 |
| 1952 | 1193111 | 3407679 | 520599 | 876795 | 1.6842 | 0.5348 |
| 1953 | 1590377 | 3557376 | 396417 | 695546 | 1.7546 | 0.3572 |
| 1954 | 641584 | 4039204 | 429694 | 826021 | 1.9223 | 0.3879 |
| 1955 | 272778 | 3488383 | 346919 | 1147841 | 3.3087 | 0.5437 |
| 1956 | 439602 | 3189831 | 299823 | 1343068 | 4.4795 | 0.6401 |
| 1957 | 804781 | 2495895 | 207840 | 792557 | 3.8133 | 0.5089 |
| 1958 | 496824 | 2164149 | 195377 | 769313 | 3.9376 | 0.5169 |
| 1959 | 683690 | 2415826 | 432489 | 744607 | 1.7217 | 0.5596 |
| 1960 | 789653 | 2050805 | 383479 | 622042 | 1.6221 | 0.4789 |
| 1961 | 916842 | 2137149 | 404228 | 783221 | 1.9376 | 0.6348 |
| 1962 | 728338 | 1957006 | 311678 | 909266 | 2.9173 | 0.7576 |
| 1963 | 472064 | 1747579 | 208207 | 776337 | 3.7287 | 0.9866 |
| 1964 | 338678 | 1374529 | 186570 | 437695 | 2.346 | 0.6789 |
| 1965 | 776941 | 1440693 | 102315 | 444930 | 4.3486 | 0.5533 |
| 1966 | 1582560 | 2198418 | 120722 | 483711 | 4.0068 | 0.5302 |
| 1967 | 1295416 | 2852164 | 129784 | 572605 | 4.412 | 0.5439 |
| 1968 | 164955 | 3387455 | 227215 | 1074084 | 4.7272 | 0.5704 |
| 1969 | 112039 | 2805591 | 151870 | 1197226 | 7.8832 | 0.8292 |
| 1970 | 197105 | 2057698 | 224482 | 933246 | 4.1573 | 0.7493 |
| 1971 | 404774 | 1610969 | 311662 | 689048 | 2.2109 | 0.5956 |
| 1972 | 1015319 | 1621485 | 346511 | 565254 | 1.6313 | 0.6928 |
| 1973 | 1818949 | 2401955 | 332913 | 792685 | 2.3811 | 0.602 |
| 1974 | 523916 | 2236387 | 164491 | 1102433 | 6.7021 | 0.5633 |
| 1975 | 621616 | 2037430 | 142028 | 829377 | 5.8395 | 0.6595 |
| 1976 | 613942 | 1931396 | 171238 | 867463 | 5.0658 | 0.6457 |
| 1977 | 348054 | 1950748 | 341385 | 905301 | 2.6518 | 0.8379 |
| 1978 | 638490 | 1576565 | 241536 | 698715 | 2.8928 | 0.9406 |
| 1979 | 198490 | 1114381 | 174699 | 440538 | 2.5217 | 0.7264 |
| 1980 | 137735 | 863862 | 108253 | 380434 | 3.5143 | 0.7241 |
| 1981 | 150868 | 983658 | 166926 | 399038 | 2.3905 | 0.8632 |
| 1982 | 151830 | 750871 | 326133 | 363730 | 1.1153 | 0.7583 |
| 1983 | 166831 | 738675 | 327181 | 289992 | 0.8863 | 0.756 |
| 1984 | 397595 | 817497 | 251087 | 277651 | 1.1058 | 0.9161 |
| 1985 | 523470 | 957429 | 193856 | 307920 | 1.5884 | 0.7038 |
| 1986 | 930301 | 1260698 | 170729 | 430113 | 2.5193 | 0.8649 |
| 1987 | 270553 | 1122943 | 121243 | 523071 | 4.3142 | 0.951 |
| 1988 | 202921 | 915093 | 202589 | 434939 | 2.1469 | 0.9743 |
| 1989 | 172782 | 890360 | 234716 | 332481 | 1.4165 | 0.6602 |
| 1990 | 242750 | 962675 | 316418 | 212000 | 0.67 | 0.271 |
| 1991 | 408186 | 1559853 | 704744 | 319158 | 0.4529 | 0.321 |
| 1992 | 700405 | 1901909 | 887537 | 513234 | 0.5783 | 0.455 |
| 1993 | 759326 | 2295839 | 774584 | 581611 | 0.7509 | 0.553 |
| 1994 | 516667 | 2023164 | 612352 | 771086 | 1.2592 | 0.8687 |
| 1995 | 306766 | 1689840 | 528033 | 739999 | 1.4014 | 0.7885 |
| 1996 | 257279 | 1597385 | 570574 | 732228 | 1.2833 | 0.7011 |
| 1997 | 491583 | 1473560 | 588529 | 762403 | 1.2954 | 1.0348 |
| 1998 | 600510 | 1159155 | 385638 | 592624 | 1.5367 | 0.9181 |
| 1999 | 469731 | 1078374 | 292763 | 484910 | 1.6563 | 0.9892 |
| 2000 | 553207 | 1074915 | 239594 | 414868 | 1.7315 | 0.8554 |
| 2001 | 483301 | 1353656 | 353526 | 426471 | 1.2063 | 0.7218 |
| 2002 | 401878 | 1527198 | 496236 | 535045 | 1.0782 | 0.6841 |
| 2003 | 649338 | 1605435 | 547157 | 551990 | 1.0088 | 0.5441 |
| 2004 | 275602 | 1564189 | 654240 | 606445 | 0.9269 | 0.6868 |
| 2005 | 489080 | 1549531 | 606000 | 641276 | 1.0582 | 0.7179 |
| 2006 | 549066 | 1540896 | 595278 | 537642 | 0.9032 | 0.6063 |
| 2007 | 776167 | 1774881 | 649155 | 486883 | 0.75 | 0.3774 |
| 2008 | 689405 | 2200287 | 703312 | 464171 | 0.66 | 0.3339 |
| 2009 | 556084 | 2607202 | 1069635 | 523430 | 0.4894 | 0.2762 |
|  |  |  |  |  |  |  |
| Arith. |  |  |  |  |  |  |
| Mean | 575237 | 1994586 | 412423 | 651240 | 2.253 | 0.632 |
| 0 Units | (Thousands) | (Tonnes) | (Tonnes) | (Tonnes) |  |  |
| 1 |  |  |  |  |  |  |

Table 3.27. Northeast arctic cod. Input for the short-term prediction
MFDP version 1a
Run: out-pa
Time and date: 20:53 26.04.2010
Fbar age range: 5-10

| 2010 |  |  | M |  | Mat | PF |  | PM |  | SWt |  | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 384000 |  | 0.3335 |  | 0 |  | 0 |  | 0 | 0.257 | 0.0155 | 0.861 |
|  | 4 | 423422 |  | 0.227 |  | 0.003 |  | 0 |  | 0 | 0.589 | 0.0806 | 1.194 |
|  | 5 | 417190 |  | 0.2051 |  | 0.045 |  | 0 |  | 0 | 1.183 | 0.1839 | 1.735 |
|  | 6 | 296705 |  | 0.2 |  | 0.323 |  | 0 |  | 0 | 2.052 | 0.2315 | 2.514 |
|  | 7 | 144048 |  | 0.2 |  | 0.573 |  | 0 |  | 0 | 3.181 | 0.259 | 3.509 |
|  | 8 | 71781 |  | 0.2 |  | 0.838 |  | 0 |  | 0 | 4.8 | 0.3414 | 5.234 |
|  | 9 | 23961 |  | 0.2 |  | 0.927 |  | 0 |  | 0 | 6.759 | 0.3661 | 6.83 |
|  | 10 | 17527 |  | 0.2 |  | 0.97 |  | 0 |  | 0 | 7.859 | 0.2753 | 7.912 |
|  | 11 | 5182 |  | 0.2 |  | 0.974 |  | 0 |  | 0 | 10.008 | 0.3002 | 9.492 |
|  | 12 | 1944 |  | 0.2 |  | 0.986 |  | 0 |  | 0 | 12.731 | 0.4831 | 10.151 |
|  | 13 | 1111 |  | 0.2 |  | 1 |  | 0 |  | 0 | 14.311 | 0.4831 | 10.031 |
| 2011 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Age | N |  | M |  | Mat |  | PF |  | PM |  |  | Sel | CWt |
|  | 3 | 465000 |  | 0.3335 |  | 0 |  | 0 |  | 0 | 0.246 | 0.0155 | 0.861 |
|  | 4 |  |  | 0.227 |  | 0.002 |  | 0 |  | 0 | 0.643 | 0.0806 | 1.307 |
|  | 5 |  |  | 0.2051 |  | 0.061 |  | 0 |  | 0 | 1.196 | 0.1839 | 1.756 |
|  | 6 |  |  | 0.2 |  | 0.326 |  | 0 |  | 0 | 2.092 | 0.2315 | 2.514 |
|  | 7 |  |  | 0.2 |  | 0.622 |  | 0 |  | 0 | 3.181 | 0.259 | 3.604 |
|  | 8 |  |  | 0.2 |  | 0.852 |  | 0 |  | 0 | 4.684 | 0.3414 | 4.879 |
|  | 9 |  |  | 0.2 |  | 0.944 |  | 0 |  | 0 | 6.563 | 0.3661 | 6.718 |
|  | 10 |  |  | 0.2 |  | 0.987 |  | 0 |  | 0 | 8.707 | 0.2753 | 8.315 |
|  | 11. |  |  | 0.2 |  | 0.985 |  | 0 |  | 0 | 8.405 | 0.3002 | 9.397 |
|  | 12 |  |  | 0.2 |  | 0.995 |  | 0 |  | 0 | 12.731 | 0.4831 | 10.977 |
|  | 13 |  |  | 0.2 |  | 1 |  | 0 |  | 0 | 14.311 | 0.4831 | 11.636 |
| 2012 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Age | N |  | M |  | Mat |  | PF |  | PM |  |  | Sel | CWt |
|  | 3 | 484000 |  | 0.3335 |  | 0 |  | 0 |  | 0 | 0.262 | 0.0155 | 0.861 |
|  | 4 |  |  | 0.227 |  | 0.002 |  | 0 |  | 0 | 0.632 | 0.0806 | 1.307 |
|  | 5 |  |  | 0.2051 |  | 0.061 |  | 0 |  | 0 | 1.25 | 0.1839 | 1.87 |
|  | 6 |  |  | 0.2 |  | 0.326 |  | 0 |  | 0 | 2.105 | 0.2315 | 2.535 |
|  | 7 |  |  | 0.2 |  | 0.622 |  | 0 |  | 0 | 3.22 | 0.259 | 3.604 |
|  | 8 |  |  | 0.2 |  | 0.852 |  | 0 |  | 0 | 4.685 | 0.3414 | 4.974 |
|  | 9 |  |  | 0.2 |  | 0.944 |  | 0 |  | 0 | 6.448 | 0.3661 | 6.363 |
|  | 10 |  |  | 0.2 |  | 0.987 |  | 0 |  | 0 | 8.511 | 0.2753 | 8.202 |
|  | 11 |  |  | 0.2 |  | 0.985 |  | 0 |  | 0 | 9.253 | 0.3002 | 9.799 |
|  | 12 |  |  | 0.2 |  | 0.995 |  | 0 |  | 0 | 12.731 | 0.4831 | 10.881 |
|  | 13 |  |  | 0.2 |  | 1 |  | 0 |  | 0 | 14.311 | 0.4831 | 12.461 |
| 2013 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Age | N |  | M |  | Mat |  | PF |  | PM |  |  | Sel | CWt |
|  | 3 | 589000 |  | 0.3335 |  | 0 |  | 0 |  | 0 | 0.262 | 0.0155 | 0.861 |
|  | 4 |  |  | 0.227 |  | 0.002 |  | 0 |  | 0 | 0.632 | 0.0806 | 1.307 |
|  | 5 |  |  | 0.2051 |  | 0.061 |  | 0 |  | 0 | 1.25 | 0.1839 | 1.87 |
|  | 6 |  |  | 0.2 |  | 0.326 |  | 0 |  | 0 | 2.105 | 0.2315 | 2.535 |
|  | 7 |  |  | 0.2 |  | 0.622 |  | 0 |  | 0 | 3.22 | 0.259 | 3.604 |
|  | 8 |  |  | 0.2 |  | 0.852 |  | 0 |  | 0 | 4.685 | 0.3414 | 4.974 |
|  | 9 |  |  | 0.2 |  | 0.944 |  | 0 |  | 0 | 6.448 | 0.3661 | 6.363 |
|  | 10 |  |  | 0.2 |  | 0.987 |  | 0 |  | 0 | 8.511 | 0.2753 | 8.202 |
|  | 11. |  |  | 0.2 |  | 0.985 |  | 0 |  | 0 | 9.253 | 0.3002 | 9.799 |
|  | 12 |  |  | 0.2 |  | 0.995 |  | 0 |  | 0 | 12.731 | 0.4831 | 10.881 |
|  | 13 |  |  | 0.2 |  | 1 |  | 0 |  | 0 | 14.311 | 0.4831 | 12.461 |

Input units are thousands and kg - output in tonnes

Table 3.28. Northeast arctic cod. Management option table.
MFDP version 1a
Run: out-pa
preMFDP Index file 25.04.2005
Time and date: 09:59 27.04.2010
Fbar age range: 5-10

| 2010 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar |  | Landings |
| 2645431 | 1145460 | 1 | 0.2762 | 592522 |  |


| 2011 |  |  | 2012 |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar |  | Landings |  |  |  | Biomass |  | SSB |
| 2838054 | 1488265 | 0 | 0 | 0 | 3668732 | 2294694 |  |  |  |  |  |
| . | 1488265 | 0.1 | 0.0276 | 73371 | 3585217 | 2230437 |  |  |  |  |  |
| . | 1488265 | 0.2 | 0.0552 | 144828 | 3503985 | 2168071 |  |  |  |  |  |
| . | 1488265 | 0.3 | 0.0829 | 214427 | 3424970 | 2107538 |  |  |  |  |  |
| . | 1488265 | 0.4 | 0.1105 | 282221 | 3348106 | 2048782 |  |  |  |  |  |
| . | 1488265 | 0.5 | 0.1381 | 348260 | 3273331 | 1991748 |  |  |  |  |  |
| . | 1488265 | 0.6 | 0.1657 | 412597 | 3200583 | 1936382 |  |  |  |  |  |
| . | 1488265 | 0.7 | 0.1933 | 475278 | 3129803 | 1882635 |  |  |  |  |  |
| . | 1488265 | 0.8 | 0.221 | 536352 | 3060935 | 1830456 |  |  |  |  |  |
| . | 1488265 | 0.9 | 0.2486 | 595863 | 2993922 | 1779797 |  |  |  |  |  |
| . | 1488265 | 1 | 0.2762 | 653857 | 2928710 | 1730613 |  |  |  |  |  |
| . | 1488265 | 1.1 | 0.3038 | 710375 | 2865248 | 1682857 |  |  |  |  |  |
| . | 1488265 | 1.2 | 0.3314 | 765460 | 2803485 | 1636486 |  |  |  |  |  |
| . | 1488265 | 1.3 | 0.3591 | 819152 | 2743372 | 1591460 |  |  |  |  |  |
| . | 1488265 | 1.4 | 0.3867 | 871491 | 2684861 | 1547735 |  |  |  |  |  |
| . | 1488265 | 1.5 | 0.4143 | 922513 | 2627906 | 1505274 |  |  |  |  |  |
| . | 1488265 | 1.6 | 0.4419 | 972256 | 2572463 | 1464038 |  |  |  |  |  |
| . | 1488265 | 1.7 | 0.4695 | 1020756 | 2518488 | 1423989 |  |  |  |  |  |
| . | 1488265 | 1.8 | 0.4972 | 1068047 | 2465939 | 1385092 |  |  |  |  |  |
| . | 1488265 | 1.9 | 0.5248 | 1114163 | 2414776 | 1347313 |  |  |  |  |  |
| . | 1488265 | 2 | 0.5524 | 1159137 | 2364959 | 1310617 |  |  |  |  |  |

Input units are thousands and kg - output in tonnes

Table 3.29a. Northeast arctic cod. Detailed prediction output assuming Fpa in 2011-2013
MFDP version 1a
Run: out-r
Time and date: 19:58 27.04.2010
Fbar age range: 5-10

| Year: Age | 2010 F multiplie |  |  |  | 1 Fbar: | 0.2762 | SSNos(Jar SSB(Jan) |  | SSNos(ST | SSB(ST) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | CatchNos | Yield | StockNos | Biomass |  |  |  |  |
|  | 3 | 0.0155 | 5024 | 4326 | 384000 | 98688 | 0 | 0 | 0 | 0 |
|  | 4 | 0.0806 | 29378 | 35078 | 423422 | 249396 | 1270 | 748 | 1270 | 748 |
|  | 5 | 0.1839 | 63559 | 110276 | 417190 | 493536 | 18774 | 22209 | 18774 | 22209 |
|  | 6 | 0.2315 | 55788 | 140251 | 296705 | 608839 | 95836 | 196655 | 95836 | 196655 |
|  | 7 | 0.259 | 29919 | 104985 | 144048 | 458217 | 82540 | 262558 | 82540 | 262558 |
|  | 8 | 0.3414 | 18923 | 99045 | 71781 | 344549 | 60152 | 288732 | 60152 | 288732 |
|  | 9 | 0.3661 | 6698 | 45749 | 23961 | 161952 | 22212 | 150130 | 22212 | 150130 |
|  | 10 | 0.2753 | 3840 | 30386 | - 17527 | 137745 | 17001 | 133612 | 17001 | 133612 |
|  | 11 | 0.3002 | 1224 | 11619 | - 5182 | 51861 | 5047 | 50513 | 5047 | 50513 |
|  | 12 | 0.4831 | 680 | 6907 | 1944 | 24749 | 1917 | 24403 | 1917 | 24403 |
|  | 13 | 0.4831 | 389 | 3901 | 1111 | 15900 | 1111 | 15900 | 1111 | 15900 |
| Total |  |  | 215424 | 592522 | 1786871 | 2645431 | 305860 | 1145460 | 305860 | 1145460 |
| Year: |  | 2011 | F multiplie | 1.4482 | Fbar: | 0.4 |  |  |  |  |
| Age |  |  | CatchNos | Yield | StockNos | Biomass | SSNos(Jar | SB(Jan) | SSNos(ST | SSB(ST) |
|  | 3 | 0.0224 | 8782 | 7562 | 2465000 | 114390 | 0 | 0 | 0 | 0 |
|  | 4 | 0.1167 | 26756 | 34970 | 270871 | 174170 | 542 | 348 | 542 | 348 |
|  | 5 | 0.2663 | 66106 | 116082 | 311304 | 372319 | 18990 | 22711 | 18990 | 22711 |
|  | 6 | 0.3353 | 73403 | 184536 | - 282744 | 591500 | 92175 | 192829 | 92175 | 192829 |
|  | 7 | 0.3751 | 54972 | 198121 | 192720 | 613042 | 119872 | 381312 | 119872 | 381312 |
|  | 8 | 0.4944 | 32446 | 158303 | 91026 | 426366 | 77554 | 363264 | 77554 | 363264 |
|  | 9 | 0.5302 | 15717 | 105584 | 41772 | 274148 | 39433 | 258796 | 39433 | 258796 |
|  | 10 | 0.3987 | 4081 | 33932 | - 13604 | 118446 | 13427 | 116906 | 13427 | 116906 |
|  | 11 | 0.4347 | 3507 | 32957 | 10897 | 91585 | 10733 | 90211 | 10733 | 90211 |
|  | 12 | 0.6996 | 1450 | 15915 | - 3142 | 40006 | 3127 | 39806 | 3127 | 39806 |
|  | 13 | 0.6996 | 712 | 8283 | 1543 | 22081 | 1543 | 22081 | 1543 | 22081 |
| Total |  |  | 287932 | 896245 | 1684622 | 2838054 | 377394 | 1488265 | 377394 | 1488265 |




Input units are thousands and kg - output in tonnes

Table 3.29b. Northeast arctic cod. Detailed prediction output assuming HCR in 2011 and Fpa in 2012-2013 MFDP version 1a
Run: our-r30
Time and date: 20:26 27.04.2010
Fbar age range: 5-10

| Year: | 2010 F multiplie |  |  | 1 Fbar: |  | 0.2762 | SSNos(Jar SSB(Jan) |  | SSNos(ST | SSB(ST) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  |  | CatchNos | Yield | StockNos | Biomass |  |  |  |  |
|  | 3 | 0.0155 | 5024 | 4326 | 384000 | 98688 | 0 | 0 | 0 | 0 |
|  | 4 | 0.0806 | 29378 | 35078 | 423422 | 249396 | 1270 | 748 | 1270 | 748 |
|  | 5 | 0.1839 | 63559 | 110276 | 417190 | 493536 | 18774 | 22209 | 18774 | 22209 |
|  | 6 | 0.2315 | 55788 | 140251 | 296705 | 608839 | 95836 | 196655 | 95836 | 196655 |
|  | 7 | 0.259 | 29919 | 104985 | 144048 | 458217 | 82540 | 262558 | 82540 | 262558 |
|  | 8 | 0.3414 | 18923 | 99045 | 71781 | 344549 | 60152 | 288732 | 60152 | 288732 |
|  | 9 | 0.3661 | 6698 | 45749 | 23961 | 161952 | 22212 | 150130 | 22212 | 150130 |
|  | 10 | 0.2753 | 3840 | 30386 | 17527 | 137745 | 17001 | 133612 | 17001 | 133612 |
|  | 11 | 0.3002 | 1224 | 11619 | 5182 | 51861 | 5047 | 50513 | 5047 | 50513 |
|  | 12 | 0.4831 | 680 | 6907 | 1944 | 24749 | 1917 | 24403 | 1917 | 24403 |
|  | 13 | 0.4831 | 389 | 3901 | 1111 | 15900 | 1111 | 15900 | 1111 | 15900 |
| Total |  |  | 215424 | 592522 | 1786871 | 2645431 | 305860 | 1145460 | 305860 | 1145460 |
| Year: Age | F 2011 |  | F multiplie | 1.0862 | Fbar: | 0.3 |  |  |  |  |
|  |  |  | CatchNos | Yield | StockNos | Biomass | SSNos(Jar SSB(Jan) |  | SSNos(ST SSB(ST) |  |
|  | 3 | 0.0168 | 6604 | 5686 | 465000 | 114390 | 0 | 0 | 0 | 0 |
|  | 4 | 0.0875 | 20347 | 26593 | 270871 | 174170 | 542 | 348 | 542 | 348 |
|  | 5 | 0.1998 | 51136 | 89795 | 311304 | 372319 | 18990 | 22711 | 18990 | 22711 |
|  | 6 | 0.2515 | 57214 | 143836 | 282744 | 591500 | 92175 | 192829 | 92175 | 192829 |
|  | 7 | 0.2813 | 43033 | 155091 | 192720 | 613042 | 119872 | 381312 | 119872 | 381312 |
|  | 8 | 0.3708 | 25720 | 125486 | 91026 | 426366 | 77554 | 363264 | 77554 | 363264 |
|  | 9 | 0.3977 | 12504 | 84003 | 41772 | 274148 | 39433 | 258796 | 39433 | 258796 |
|  | 10 | 0.299 | 3203 | 26629 | 13604 | 118446 | 13427 | 116906 | 13427 | 116906 |
|  | 11 | 0.3261 | 2763 | 25963 | 10897 | 91585 | 10733 | 90211 | 10733 | 90211 |
|  | 12 | 0.5247 | 1173 | 12876 | 3142 | 40006 | 3127 | 39806 | 3127 | 39806 |
|  | 13 | 0.5247 | 576 | 6702 | 1543 | 22081 | 1543 | 22081 | 1543 | 22081 |
| Total |  |  | 224273 | 702662 | 1684622 | 2838054 | 377394 | 1488265 | 377394 | 1488265 |
| Year: Age | F |  | F multiplie | 1.4482 | bar: | 0.4 |  |  |  |  |
|  |  |  | CatchNos | Yield | StockNos | Biomass | SSNos(Jar SSB(Jan) |  | SSNos(ST SSB(ST) |  |
|  | 3 | 0.0224 | 9141 | 7871 | 484000 | 126808 | 0 | 0 | 0 | 0 |
|  | 4 | 0.1167 | 32357 | 42290 | 327570 | 207024 | 655 | 414 | 655 | 414 |
|  | 5 | 0.2663 | 41996 | 78533 | 197768 | 247210 | 12064 | 15080 | 12064 | 15080 |
|  | 6 | 0.3353 | 53911 | 136666 | 207663 | 437131 | 67698 | 142505 | 67698 | 142505 |
|  | 7 | 0.3751 | 51351 | 185068 | 180023 | 579675 | 111975 | 360558 | 111975 | 360558 |
|  | 8 | 0.4944 | 42450 | 211149 | 119094 | 557955 | 101468 | 475378 | 101468 | 475378 |
|  | 9 | 0.5302 | 19352 | 123139 | 51435 | 331653 | 48555 | 313080 | 48555 | 313080 |
|  | 10 | 0.3987 | 6893 | 56538 | 22979 | 195571 | 22680 | 193028 | 22680 | 193028 |
|  | 11 | 0.4347 | 2658 | 26048 | 8259 | 76420 | 8135 | 75274 | 8135 | 75274 |
|  | 12 | 0.6996 | 2971 | 32326 | 6439 | 81974 | 6407 | 81564 | 6407 | 81564 |
|  | 13 | 0.6996 | 1047 | 13050 | 2270 | 32483 | 2270 | 32483 | 2270 | 32483 |
| Total |  |  | 264129 | 912677 | 1607499 | 2873904 | 381906 | 1689364 | 381906 | 1689364 |
| Year: Age | F |  | F multiplie | 1.4482 | Fbar: | 0.4 |  |  |  |  |
|  |  |  | CatchNos Yield |  | StockNos | Biomass | SSNos(Jar SSB(Jan) |  | SSNos(ST SSB(ST) |  |
|  |  |  | 11332 | 9757 | 600000 | 157200 | 0 | 0 | 0 | 0 |
|  | 4 | 0.1167 | 33490 | 43772 | 339047 | 214278 | 678 | 429 | 678 | 429 |
|  | 5 | 0.2663 | 49327 | 92241 | 232288 | 290360 | 14170 | 17712 | 14170 | 17712 |
|  | 6 | 0.3353 | 32044 | 81230 | 123430 | 259819 | 40238 | 84701 | 40238 | 84701 |
|  | 7 | 0.3751 | 34683 | 124998 | 121590 | 391521 | 75629 | 243526 | 75629 | 243526 |
|  | 8 | 0.4944 | 36105 | 179586 | 101292 | 474551 | 86300 | 404317 | 86300 | 404317 |
|  | 9 | 0.5302 | 22376 | 142379 | 59471 | 383472 | 56141 | 361998 | 56141 | 361998 |
|  | 10 | 0.3987 | 7434 | 60976 | 24782 | 210923 | 24460 | 208181 | 24460 | 208181 |
|  | 11 | 0.4347 | 4064 | 39826 | 12627 | 116842 | 12438 | 115089 | 12438 | 115089 |
|  | 12 | 0.6996 | 2020 | 21978 | 4378 | 55734 | 4356 | 55455 | 4356 | 55455 |
|  | 13 | 0.6996 | 1634 | 20364 | 3542 | 50690 | 3542 | 50690 | 3542 | 50690 |
| Total |  |  | 234510 | 817108 | 1622447 | 2605389 | 317953 | 1542098 | 317953 | 1542098 |

Table 3.30. North East arctic cod. Stock numbers at age (in thousands) estimated by VPA including discard estimates, and \% increase in stock numbers relative to a VPA without discards. From Dingsør (2001). The discard numbers applied correspond to method II (1946-1982) and IIIb (19831998) mentioned in Dingsør (2001).

| Year | Estimated stock numbers (thousands) |  |  | Percent increase |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 3 | Age 4 | Age 5 | Age 3 | Age 4 | Age 5 |
| 1946 | 875346 | 602579 | 407163 | 20 \% | 4 \% | 1 \% |
| 1947 | 531993 | 676806 | 465099 | 27 \% | 14 \% | 0 \% |
| 1948 | 570356 | 392309 | 497476 | 29 \% | 14 \% | $5 \%$ |
| 1949 | 589367 | 416668 | 285459 | 26 \% | 16 \% | $3 \%$ |
| 1950 | 799732 | 414016 | 291200 | 13 \% | $9 \%$ | $1 \%$ |
| 1951 | 1235322 | 586054 | 302346 | 14 \% | $2 \%$ | 0 \% |
| 1952 | 1388731 | 889509 | 401768 | 17 \% | $3 \%$ | 0 \% |
| 1953 | 1801114 | 975004 | 600908 | 13 \% | $2 \%$ | 0 \% |
| 1954 | 830653 | 1321053 | 684303 | 29 \% | $5 \%$ | 0 \% |
| 1955 | 381489 | 615696 | 907875 | 40 \% | 19 \% | $2 \%$ |
| 1956 | 567555 | 274235 | 399344 | 29 \% | 25 \% | $3 \%$ |
| 1957 | 914850 | 387496 | 161710 | 14 \% | $10 \%$ | $2 \%$ |
| 1958 | 552600 | 672221 | 262135 | 11 \% | 4 \% | $2 \%$ |
| 1959 | 757567 | 391906 | 406694 | 11 \% | $3 \%$ | 0 \% |
| 1960 | 855470 | 534350 | 240047 | 8 \% | $1 \%$ | $0 \%$ |
| 1961 | 1041570 | 620707 | 347043 | 13 \% | 1 \% | 0 \% |
| 1962 | 894728 | 739196 | 382556 | 23 \% | $4 \%$ | $0 \%$ |
| 1963 | 551938 | 614025 | 429068 | 17 \% | $10 \%$ | $0 \%$ |
| 1964 | 389151 | 396165 | 361790 | 15 \% | $5 \%$ | $0 \%$ |
| 1965 | 845469 | 293844 | 266134 | 9 \% | 8 \% | 0 \% |
| 1966 | 1618188 | 647435 | 203168 | 2 \% | $4 \%$ | $2 \%$ |
| 1967 | 1404569 | 1249506 | 465035 | $9 \%$ | $0 \%$ | $1 \%$ |
| 1968 | 210875 | 1088071 | 876095 | 24 \% | $6 \%$ | $0 \%$ |
| 1969 | 143791 | 155947 | 699033 | 28 \% | $15 \%$ | $2 \%$ |
| 1970 | 222635 | 104415 | 92541 | 13 \% | 17 \% | $4 \%$ |
| 1971 | 462474 | 164397 | 65112 | 14 \% | $6 \%$ | $2 \%$ |
| 1972 | 1221559 | 358357 | 115892 | 20 \% | 10 \% | 1 \% |
| 1973 | 1858123 | 947409 | 249400 | 2 \% | 19 \% | 11 \% |
| 1974 | 598555 | 1246499 | 583612 | 14 \% | $2 \%$ | $9 \%$ |
| 1975 | 654442 | 382692 | 627793 | 5 \% | 10 \% | $3 \%$ |
| 1976 | 622230 | 477390 | 233608 | 1 \% | $2 \%$ | $1 \%$ |
| 1977 | 397826 | 426386 | 280645 | 14 \% | 0 \% | 0 \% |
| 1978 | 653256 | 277410 | 198204 | 2 \% | $11 \%$ | 0 \% |
| 1979 | 225935 | 460104 | 164243 | 14 \% | $2 \%$ | $1 \%$ |
| 1980 | 152937 | 171954 | 300312 | 11 \% | $11 \%$ | $0 \%$ |
| 1981 | 161752 | 116964 | 116337 | 7 \% | $7 \%$ | 4 \% |
| 1982 | 151642 | 125307 | 81780 | 0 \% | 4 \% | $1 \%$ |
| 1983 | 166310 | 115423 | 82423 | 0 \% | -1 \% | $3 \%$ |
| 1984 | 408525 | 133333 | 77728 | $3 \%$ | 0 \% | 0 \% |
| 1985 | 543828 | 324072 | 96327 | 4 \% | $2 \%$ | 0 \% |
| 1986 | 1114252 | 412683 | 219993 | 7 \% | $2 \%$ | 0 \% |
| 1987 | 307425 | 767656 | 268642 | 7 \% | $4 \%$ | 0 \% |
| 1988 | 222819 | 215720 | 490161 | $9 \%$ | $3 \%$ | $2 \%$ |
| 1989 | 180066 | 166955 | 151576 | 4 \% | $6 \%$ | 0 \% |
| 1990 | 249968 | 139922 | 114006 | $3 \%$ | $2 \%$ | $1 \%$ |
| 1991 | 418955 | 200700 | 105559 | 2 \% | $2 \%$ | 0 \% |
| 1992 | 748962 | 333517 | 151973 | 4 \% | $1 \%$ | 0 \% |
| 1993 | 1002933 | 576112 | 238980 | 10 \% | $2 \%$ | 0 \% |
| 1994 | 896184 | 744062 | 420039 | $9 \%$ | 8 \% | 0 \% |
| 1995 | 733664 | 584808 | 476048 | 10 \% | $6 \%$ | $3 \%$ |
| 1996 | 467093 | 341918 | 344124 | 3 \% | $7 \%$ | $3 \%$ |
| 1997 | 765234 | 238202 | 193102 | 3 \% | $0 \%$ | 4 \% |
| 1998 | 836301 | 429147 | 144629 | $2 \%$ | $1 \%$ | -1 \% |

Table 3.31. Northeast Arctic cod. Number (thousands) of cod by age groups taken as by-catch in the Norwegian shrimp fishery (1984-2006)

| Age \ Year | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 322 | 4537 | 28 | 1408 | 259 | 717 | 2971 | 11651 |
| 1 | 4913 | 19437 | 2339 | 3259 | 1719 | 668 | 13731 | 34450 |
| 2 | 1624 | 49334 | 6952 | 1961 | 1534 | 418 | 1518 | 2759 |
| 3 | 1073 | 2720 | 5245 | 499 | 1380 | 694 | 1019 | 87 |
| 4 | 2200 | 1891 | 716 | 2210 | 1882 | 2096 | 403 | 64 |
| 5 | 161 | 9306 | 737 | 1715 | 1124 | 2281 | 909 | 33 |
| 6 | 89 | 6374 | 520 | 411 | 269 | 1135 | 2913 | 293 |
| 7 | 144 | 266 | 92 | 79 | 186 | 184 | 1434 | 1138 |
| 8 | 38 | 1 | 93 | 28 | 178 | 13 | 185 | 316 |
| 9 | 1 | 2 | 165 | 6 | 1 | 0 | 3 | 29 |
| 10 | 0 | 3 | 88 | 1 | 0 | 0 | 9 | 0 |
| 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total('000) | 10564 | 93872 | 16976 | 11576 | 8532 | 8206 | 25095 | 50819 |
| Age \ Year | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 6486 | 604 | 1042 | 1138 | 519 | 896 | 506 | 651 |
| 1 | 5236 | 6702 | 1628 | 1896 | 9084 | 17157 | 40314 | 7155 |
| 2 | 2922 | 4032 | 410 | 99 | 359 | 1805 | 5248 | 245 |
| 3 | 242 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total('000) | 14886 | 11339 | 3080 | 3133 | 9962 | 19858 | 46068 | 8052 |


| Age $\backslash$ Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 66 | 1188 | 478 | 4253 | 713 | 945 | 1355 |
| 1 | 1572 | 7187 | 293 | 8805 | 1014 | 3411 | 2597 |
| 2 | 3152 | 1348 | 893 | 96 | 323 | 1628 | 218 |
| 3 | 218 | 0 | 190 | 0 | 0 | 0 | 0 |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total('000) | 5007 | 9723 | 1854 | 13154 | 2051 | 5984 | 4170 |





Figure 3.1. ICES Standard plots for Northeast Arctic cod (sub-area I and II)




Figure. 3.1. Continued. ICES Standard plots for Northeast Arctic cod (sub-area I and II)


Figure 3.2. Northeast Arctic cod. Log catchability residual (y-axis) by fleets for the tuning data used in xsa. Ages 3-5 in left hand panel and 6-8 in right hand panel.


Figure 3.2 continued.... Northeast Arctic cod. Log catchability residual (y-axis) by fleets for the tuning data used in xsa. Ages 9-11.


Figure 3.3. Northeast Arctic cod. Single fleet estimates (before shrinkage) of F2009 and survivors at the end of 2009 taken from xsa-diagnostics of single fleet runs. "ALL" is the estimates from the final xsa (with shrinkage, including all fleets). The Fs for ages 3-5 includes cannibalism mortality.


Figure 3.3. continued....Single fleet estimates (before shrinkage) of F2009 and survivors at the end of 2009 taken from xsa-diagnostics of single fleet runs. "ALL" is the estimate from the final xsa (with shrinkage, including all fleets).



Figure 3.4. Northeast Arctic cod. Retrospective plots with catchability dependent on stock size for ages $<6$.


Figure 3.5. Northeast Arctic cod. The profiles of the components of the TISVPA loss function for C(a,y), fleet 1 (FLT09), fleet 2 (FLT15), fleet 3 (FLT16), fleet 4 (FLT18) and Total sum.




Figure 3.6. Northeast Arctic cod. TISVPA residuals in logarithmic catch-at-age and abundance-atage for fleets 1-4.The circles on the low line correspond to +1 (yellow) or $\mathbf{- 1}$ (grey).



Figure 3.6. Continued.




Figure 3.7. Northeast Arctic cod. TISVPA retrospective runs


Figure 3.8. Northeast Arctic cod. Comparison of TISVPA-derived estimates of recruitment to catches at age 3 and to index of the same age-group obtained from Fleet 2.



Figure 3.9. Northeast Arctic cod. Comparison of XSA and TISVPA runs

|  |  |  |
| :---: | :---: | :---: |
|  |  |  |
|  |  |  |
|  |  |  |

Figure 3.10a. Northeast Arctic cod. Weight in catch predictions.


Figure 3.10b. Northeast Arctic cod. Weight in stock projections


Figure 3.11. Northeast Arctic cod. Capelin biomass and cannibalism mortality on cod age 1,2 and 3.


Figure 3.12. Northeast arctic cod. Calibrated (with intercept) bottom trawl survey estimates (connected solid circles), ICES 2010 estimates (connected open diamonds) and the 1995-2009 ICES annual assessments (unconnected symbols) of the total number of Northeast Arctic cod ages 4 through 6.


Figure 3.13 Calibrated (with intercept) bottom trawl survey estimates (connected solid diamonds), ICES 2010 estimates (connected open circles) and the 1995-2009 ICES annual assessments (unconnected symbols) of the total number of Northeast Arctic codages 7 and older.


Figure 3.14. Spawning stock biomass, stock biomass (3+) and recruitment from the 2010 Gadget run for Northeast Arctic Cod, compared with the 2009 model run.

Table A1 North-East Arctic COD. Catch per unit effort.

| Year | Sub-area $\mid$ I |  |  | Division IIb |  |  | $\begin{array}{\|l\|} \hline \text { Division IIa } \\ \hline \text { Norway }^{2} \\ \hline \end{array}$ |  | Total <br> Norway |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Norway ${ }^{2}$ | $\mathrm{UK}^{3}$ | Russia ${ }^{4}$ | Norway ${ }^{2}$ | $\mathrm{UK}^{3}$ | Russia ${ }^{4}$ |  | UK ${ }^{3}$ |  |
| 1960 | - | 0.075 | 0.42 | - | 0.105 | 0.31 | - | 0.067 |  |
| 1961 | - | 0.079 | 0.38 | - | 0.129 | 0.44 | - | 0.058 |  |
| 1962 | - | 0.092 | 0.59 | - | 0.133 | 0.74 | - | 0.066 |  |
| 1963 | - | 0.085 | 0.60 | - | 0.098 | 0.55 | - | 0.066 |  |
| 1964 | - | 0.056 | 0.37 | - | 0.092 | 0.39 | - | 0.070 |  |
| 1965 | - | 0.066 | 0.39 | - | 0.109 | 0.49 | - | 0.066 |  |
| 1966 | - | 0.074 | 0.42 | - | 0.078 | 0.19 | - | 0.067 |  |
| 1967 | - | 0.081 | 0.53 | - | 0.106 | 0.87 | - | 0.052 |  |
| 1968 | - | 0.110 | 1.09 |  | 0.173 | 1.21 | - | 0.056 |  |
| 1969 | - | 0.113 | 1.00 | - | 0.135 | 1.17 | - | 0.094 |  |
| 1970 | - | 0.100 | 0.80 | - | 0.100 | 0.80 | - | 0.066 |  |
| 1971 | - | 0.056 | 0.43 | - | 0.071 | 0.16 | - | 0.062 |  |
| 1972 | 0.90 | 0.047 | 0.34 | 0.59 | 0.051 | 0.18 | 1.08 | 0.055 |  |
| 1973 | 1.05 | 0.057 | 0.56 | 0.43 | 0.054 | 0.57 | 0.71 | 0.043 |  |
| 1974 | 1.75 | 0.079 | 0.86 | 1.94 | 0.106 | 0.77 | 0.19 | 0.028 |  |
| 1975 | 1.82 | 0.077 | 0.94 | 1.67 | 0.100 | 0.43 | 1.36 | 0.033 |  |
| 1976 | 1.69 | 0.060 | 0.84 | 1.20 | 0.081 | 0.30 | 1.69 | 0.035 |  |
| 1977 | 1.54 | 0.052 | 0.63 | 0.91 | 0.056 | 0.25 | 1.16 | 0.044 | 1.17 |
| 1978 | 1.37 | 0.062 | 0.52 | 0.56 | 0.044 | 0.08 | 1.12 | 0.037 | 0.94 |
| 1979 | 0.85 | 0.046 | 0.43 | 0.62 | . | 0.06 | 1.06 | 0.042 | 0.85 |
| 1980 | 1.47 | , | 0.49 | 0.41 |  | 0.16 | 1.27 |  | 1.23 |
|  |  |  |  |  | Spain ${ }^{5}$ |  |  | $\text { Russia }^{4}$ |  |
| 1981 | 1.42 | - | 0.41 | (0.96) | Spais | 0.07 | 1.02 | 0.35 | 1.21 |
| 1982 | 1.30 | - | 0.35 | - | 0.86 | 0.26 | 1.01 | 0.34 | 1.09 |
| 1983 | 1.58 | - | 0.31 | (1.31) | 0.92 | 0.36 | 1.05 | 0.38 | 1.11 |
| 1984 | 1.40 | - | 0.45 | 1.20 | 0.78 | 0.35 | 0.73 | 0.27 | 0.96 |
| 1985 | 1.86 | - | 1.04 | 1.51 | 1.37 | 0.50 | 0.90 | 0.39 | 1.29 |
| 1986 | 1.97 | - | 1.00 | 2.39 | 1.73 | 0.84 | 1.36 | 1.14 | 1.70 |
| 1987 | 1.77 | - | 0.97 | 2.00 | 1.82 | 1.05 | 1.73 | 0.67 | 1.77 |
| 1988 | 1.58 | - | 0.66 | 1.61 | (1.36) | 0.54 | 0.97 | 0.55 | 1.03 |
| 1989 | 1.49 | - | 0.71 | 0.41 | 2.70 | 0.45 | 0.78 | 0.43 | 0.76 |
| 1990 | 1.35 | - | 0.70 | 0.39 | 2.69 | 0.80 | 0.38 | 0.60 | 0.49 |
| 1991 | 1.38 | - | 0.67 | 0.29 | 4.96 | 0.76 | 0.50 | 0.90 | 0.44 |
| 1992 | 2.19 | - | 0.79 | 3.06 | 2.47 | 0.23 | 0.98 | 0.65 | 1.29 |
| 1993 | 2.33 | - | 0.85 | 2.98 | 3.38 | 1.00 | 1.74 | 1.03 | 1.87 |
| 1994 | 2.50 | - | 1.01 | 2.82 | 1.44 | 1.14 | 1.27 | 0.86 | 1.59 |
| 1995 | 1.57 | - | 0.59 | 2.73 | 1.65 | 1.10 | 1.00 | 1.01 | 1.92 |
| 1996 |  |  | 0.74 |  | 1.11 | 0.85 |  | 0.99 | 1.81 |
| 1997 |  |  | 0.61 |  |  | 0.57 |  | 0.74 | 1.36 |
| 1998 |  |  | 0.37 |  |  | 0.29 |  | 0.40 | 0.83 |
| 1999 |  |  | 0.29 |  |  | 0.34 |  | 0.39 | 0.74 |
| 2000 |  |  | 0.34 |  |  | 0.37 |  | 0.53 | 0.92 |
| 2001 |  |  | 0.46 |  |  | 0.46 |  | 0.69 | 1.21 |
| 2002 |  |  | 0.58 |  |  | 0.66 |  | 0.57 | 1.35 |
| 2003 |  |  | 0.70 |  |  | 1.22 |  | 0.73 | 1.67 |
| 2004 |  |  | 0.48 |  |  | 0.78 |  | 0.84 | 1.67 |
| 2005 |  |  | 0.45 |  |  | 0.62 |  | 0.81 | 1.23 |
| 2006 |  |  | 0.49 |  |  | 0.54 |  | 0.84 | 0.88 |
| 2007 |  |  | 0.71 |  |  | 0.51 |  | 0.88 | 1.16 |
| 2008 |  |  | 0.93 |  |  | 0.79 |  | 1.21 |  |
| $2009{ }^{1}$ |  |  | 1.33 |  |  | 1.16 |  | 0.83 |  |

${ }^{1}$ Preliminary figures.
${ }^{2}$ Norwegian data - t per 1,000 tonnage*hrs fishing
${ }^{3}$ United Kingdom data - t per 100 tonnage*hrs fishing.
${ }^{4}$ Russian data - t per hr fishing
${ }^{5}$ Spanish data - t per hr fishing.

| Period | Sub-area I | Divisions IIa and Ilb |
| :--- | :---: | :---: |
| $1960-1973$ | RT | RT |
| $1974-1980$ | PST | RT |
| $1981-$ | PST | PST |

Vessel type:
RT $=$ side trawlers, $800-1000 \mathrm{HP}, \mathrm{PST}=$ stern trawlers, up to 2000 HP .

Table A2. North-east Arctic COD. Abundance indices (millions) from the Norwegian acoustic survey
in the Barents Sea in January-March. New TS and rock-hopper gear (1981-1988 back-calculated from bobbins gear). Corrected for length-dependent effective spread of trawl.

Year

| 1981 |
| :---: |
| 1982 |
| 1983 |
| 1984 |
| 1985 |
| 1986 |
| 1987 |
| 1988 |
| 1989 |
| 1990 |
| 1991 |
| 1992 |
| $1993{ }^{1}$ |
| $1994{ }^{1}$ |
| $1995{ }^{1}$ |
| $1996{ }^{1}$ |
| $1997{ }^{\text {1,2 }}$ |
| $1998{ }^{1.2}$ |
| $1999{ }^{1}$ |
| $2000{ }^{1}$ |
| $2001{ }^{1}$ |
| $2002{ }^{1}$ |
| $2003{ }^{1}$ |
| $2004{ }^{1}$ |
| $2005{ }^{1}$ |
| $2006{ }^{1}$ |
| $2007{ }^{\text {1,2 }}$ |
| $2008{ }^{1}$ |
| $2009{ }^{1}$ |
| $2010{ }^{1}$ |


| Age |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2 | 3 | 4 | 5 | 6 |  |
| 8.0 | 82.0 | 40.0 | 63.0 | 106.0 | 103.0 | 16. |
| 4.0 | 5.0 | 49.0 | 43.0 | 40.0 | 26.0 | 28. |
| 60.5 | 2.8 | 5.3 | 14.3 | 17.4 | 11.1 | 5. |
| 745.4 | 146.1 | 39.1 | 13.6 | 11.3 | 7.4 | 2. |
| 69.1 | 446.3 | 153.0 | 141.6 | 19.7 | 7.6 | 3. |
| 353.6 | 243.9 | 499.6 | 134.3 | 65.9 | 8.3 | 2. |
| 1.6 | 34.1 | 62.8 | 204.9 | 41.4 | 10.4 | 1. |
| 2.0 | 26.3 | 50.4 | 35.5 | 56.2 | 6.5 | 1. |
| 7.5 | 8.0 | 17.0 | 34.4 | 21.4 | 53.8 | 6. |
| 81.1 | 24.9 | 14.8 | 20.6 | 26.1 | 24.3 | 39. |
| 181.0 | 219.5 | 50.2 | 34.6 | 29.3 | 28.9 | 16. |
| 241.4 | 562.1 | 176.5 | 65.8 | 18.8 | 13.2 | 7. |
| 1074.0 | 494.7 | 357.2 | 191.1 | 108.2 | 20.8 | 8. |
| 858.3 | 577.2 | 349.8 | 404.5 | 193.7 | 63.6 | 12. |
| 2619.2 | 292.9 | 166.2 | 159.8 | 210.1 | 68.8 | 16. |
| 2396.0 | 339.8 | 92.9 | 70.5 | 85.8 | 74.7 | 20. |
| 1623.5 | 430.5 | 188.3 | 51.7 | 49.3 | 37.2 | 22. |
| 3401.3 | 632.9 | 427.7 | 182.6 | 42.3 | 33.5 | 26. |
| 358.3 | 304.3 | 150.0 | 96.4 | 45.1 | 10.3 | 6. |
| 154.1 | 221.4 | 245.2 | 158.9 | 142.1 | 45.4 | 9. |
| 629.9 | 63.9 | 138.2 | 171.6 | 77.3 | 39.7 | 11. |
| 18.2 | 215.5 | 69.3 | 112.2 | 102.0 | 47.0 | 18. |
| 1693.9 | 61.5 | 303.4 | 114.4 | 129.0 | 114.9 | 34. |
| 157.6 | 105.2 | 33.6 | 92.8 | 30.7 | 27.6 | 17. |
| 465.3 | 119.6 | 123.9 | 33.7 | 62.8 | 16.9 | 14. |
| 544.6 | 216.6 | 79.8 | 59.1 | 15.5 | 25.6 | 8. |
| 125.0 | 61.7 | 80.3 | 37.1 | 30.4 | 9.1 | 14. |
| 68.8 | 97.6 | 210.2 | 306.1 | 140.6 | 69.4 | 21. |
| 321.5 | 30.6 | 182.6 | 178.3 | 137.1 | 35.0 | 12. |
| 485.4 | 59.4 | 34.7 | 121.9 | 174.7 | 162.3 | 44. |
| ${ }^{1}$ Survey cove | a large |  |  |  |  |  |


| 8 | $910+$ | Total |  |
| :---: | :---: | :---: | :---: |
| 3.0 | 1.0 | 1.0 | 423.0 |
| 2.0 | + | 0.0 | 197.0 |
| 3.0 | 0.5 | 0.1 | 120.5 |
| 0.2 | 0.0 | 0.0 | 966.0 |
| 0.2 | 0.1 | 0.0 | 840.9 |
| 0.4 | 0.1 | 0.0 | 1308.2 |
| 0.2 | 0.7 | 0.0 | 357.3 |
| 0.2 | 0.0 | 0.0 | 178.4 |
| 1.0 | 0.1 | 0.1 | 150.1 |
| 2.4 | 0.1 | 0.0 | 234.1 |
| 17.3 | 0.9 | 0.0 | 578.7 |
| 4.5 | 2.8 | 0.2 | 1092.9 |
| 5.0 | 2.3 | 2.5 | 2264.0 |
| 3.7 | 1.7 | 0.9 | 2465.4 |
| 2.1 | 0.7 | 1.0 | 3537.4 |
| 2.8 | 0.3 | 0.4 | 3083.8 |
| 4.0 | 0.7 | 0.1 | 2407.5 |
| 13.6 | 1.7 | 0.3 | 4762.8 |
| 4.1 | 0.8 | 0.3 | 976.1 |
| 4.7 | 3.0 | 1.1 | 985.5 |
| 1.4 | 0.5 | 0.2 | 1134.5 |
| 3.0 | 0.4 | 0.3 | 585.9 |
| 7.7 | 1.9 | 0.5 | 2461.5 |
| 5.9 | 1.2 | 0.2 | 471.8 |
| 4.2 | 1.0 | 0.4 | 842.4 |
| 4.5 | 1.4 | 0.5 | 956.5 |
| 5.0 | 2.1 | 0.7 | 365.6 |
| 12.2 | 3.1 | 0.8 | 930.4 |
| 5.2 | 3.7 | 0.9 | 907.3 |
| 13.8 | 3.5 | 3.5 | 1103.6 |

23.0
197.0
120.5
966.0
840.9
308.2
357.3
178.4
150.1
234.1
578.7
1092.9
2264.0
2465.4
3537.4
3083.8
2407.5
4762.8
976.1
985.5
1134.5
585.9
2461.5
471.8
842.4
956.5
365.6
930.4
907.3
1103.6

Table A3. North-East Arctic COD. Abundance indices (millions) from the Norwegian bottom traw survey in the Barents Sea in January-March. Rock-hopper gear (1981-1988 back-calculated from bobbins gear). Corrected for length-dependent effective spread of trawl.

| Age |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 10+ |  |  |
| 1981 | 4.6 | 34.3 | 16.4 | 23.3 | 40 | 38.4 | 4.8 | 1 | 0.3 | 0 | 163.1 |
| 1982 | 0.8 | 2.9 | 28.3 | 27.7 | 23.6 | 15.5 | 16 | 1.4 | 0.2 | 0 | 116.4 |
| 1983 | 152.9 | 13.4 | 25.0 | 52.3 | 43.3 | 17.0 | 5.8 | 3.2 | 1.0 | 0.1 | 313.9 |
| 1984 | 2755.0 | 379.1 | 97.5 | 28.3 | 21.4 | 11.7 | 4.1 | 0.4 | 0.1 | 0.1 | 3297.7 |
| 1985 | 49.5 | 660.0 | 166.8 | 126.0 | 19.9 | 7.7 | 3.3 | 0.2 | 0.1 | 0.1 | 1033.6 |
| 1986 | 665.8 | 399.6 | 805.0 | 143.9 | 64.1 | 8.3 | 1.9 | 0.3 | 0.0 | 0.0 | 2089.1 |
| 1987 | 30.7 | 445.0 | 240.4 | 391.1 | 54.3 | 15.7 | 2.0 | 0.5 | 0.0 | 0.0 | 1179.8 |
| 1988 | 3.2 | 72.8 | 148.0 | 80.5 | 173.3 | 20.5 | 3.6 | 0.5 | 0.0 | 0.0 | 502.5 |
| 1989 | 8.2 | 15.6 | 46.4 | 75.9 | 37.8 | 90.2 | 9.8 | 0.9 | 0.1 | 0.1 | 285.0 |
| 1990 | 207.2 | 56.7 | 28.4 | 34.9 | 34.6 | 20.6 | 27.2 | 1.6 | 0.4 | 0.0 | 411.5 |
| 1991 | 460.5 | 220.1 | 45.9 | 33.7 | 25.7 | 21.5 | 12.2 | 12.7 | 0.6 | 0.0 | 832.7 |
| 1992 | 126.6 | 570.9 | 158.3 | 57.7 | 17.8 | 12.8 | 7.7 | 4.3 | 2.7 | 0.2 | 959.0 |
| $1993{ }^{1}$ | 534.5 | 420.4 | 273.9 | 140.1 | 72.5 | 15.8 | 6.2 | 3.9 | 2.2 | $2.4{ }^{\text {F }}$ | 1471.9 |
| $1994{ }^{1}$ | 1035.9 | 535.8 | 296.5 | 310.2 | 147.4 | 50.6 | 9.3 | 2.4 | 1.6 | $1.3{ }^{*}$ | 2391.0 |
| $1995{ }^{1}$ | 5253.1 | 541.5 | 274.6 | 241.4 | 255.9 | 76.7 | 18.5 | 2.4 | 0.8 | $1.1{ }^{*}$ | 6666.2 |
| $1996{ }^{1}$ | 5768.5 | 707.6 | 170.0 | 115.4 | 137.2 | 106.1 | 24.0 | 2.9 | 0.4 | $0.5{ }^{\text {² }}$ | 7032.5 |
| $1997{ }^{\text {1,2}}$ | 4815.5 | 1045.1 | 238.0 | 64.0 | 70.4 | 52.7 | 28.3 | 5.7 | 0.9 | 0.5 | 6321.1 |
| $1998{ }^{1,2}$ | 2418.5 | 643.7 | 396.0 | 181.3 | 36.5 | 25.9 | 17.8 | 8.6 | 1.0 | 0.5 | 3729.8 |
| $1999{ }^{1}$ | 484.6 | 340.1 | 211.8 | 173.2 | 58.1 | 13.4 | 6.5 | 5.1 | 1.2 | 0.4 | 1294.4 |
| $2000{ }^{1}$ | 128.8 | 248.3 | 235.2 | 132.1 | 108.3 | 26.9 | 4.3 | 2.0 | 1.2 | 0.4 | 887.5 |
| $2001{ }^{1}$ | 657.9 | 76.6 | 191.1 | 182.8 | 83.4 | 38.2 | 8.9 | 1.1 | 0.4 | 0.2 | 1240.6 |
| $2002{ }^{1}$ | 35.3 | 443.9 | 88.3 | 135.0 | 109.6 | 42.5 | 15.1 | 2.4 | 0.3 | 0.2 | 872.6 |
| $2003{ }^{1}$ | 2991.7 | 79.1 | 377.0 | 129.7 | 91.1 | 67.3 | 18.3 | 4.9 | 1.0 | 0.2 | 3760.3 |
| $2004{ }^{1}$ | 328.5 | 235.4 | 76.6 | 172.5 | 56.9 | 44.7 | 27.3 | 7.6 | $1.7{ }^{\text {F }}$ | 0.4 | 951.6 |
| $2005{ }^{1}$ | 824.3 | 224.6 | 246.9 | 62.1 | 98.1 | 24.7 | 15.5 | 4.5 | 1.1 | 0.4 | 1502.3 |
| $2006{ }^{1}$ | 862.7 | 288.4 | 118.1 | 111.5 | 28.7 | 43.7 | 10.2 | 4.9 | 1.4 | 0.6 | 1470.4 |
| $2007{ }^{1,2}$ | 485.9 | 393.9 | 367.7 | 85.0 | 62.9 | 14.8 | 17.9 | 4.8 | 1.8 | 0.7 | 1435.4 |
| $2008{ }^{1}$ | 70.4 | 95.1 | 190.2 | 333.6 | 91.0 | 47.2 | 13.0 | 8.8 | 2.0 | 0.4 | 851.7 |
| $2009{ }^{1}$ | 382.7 | 39.1 | 118.3 | 219.5 | 193.9 | 58.7 | 19.6 | 6.8 | 4.8 | $0.9{ }^{\text {F }}$ | 1044.3 |
| 2010 | 1020.2 | 104.4 | 36.0 | 106.9 | 160.8 | 140.7 | 40.0 | 11.9 | 3.5 | 2.2 | 1627.0 |
|  | rvey cover | d a large es |  |  |  |  |  |  |  |  |  |

Table A4. North East Arctic COD. Abundance at age (millions) from the Norwegian acoustic survey on the spawning grounds off Lofoten in March-April.

| Year | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ | Sum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 0.68 | 7.45 | 12.36 | 3.11 | 1.15 | 1.01 | 0.45 |  | 26.21 |
| 1986 | 2.49 | 3.30 | 5.54 | 2.71 | 0.16 |  | 0.40 | $0.08{ }^{\prime \prime}$ | 14.68 |
| 1987 | 8.77 | 7.04 | 0.23 | 2.83 | 0.04 |  | 0.03 | 0.03 | 18.97 |
| 1988 | 1.57 | 4.43 | 2.56 | 0.05 | 0.01 | 0.05 |  |  | 8.67 |
| 1989 | 0.04 | 13.20 | 9.73 | 2.20 | 0.38 | 0.12 |  | 0.06 | 25.73 |
| 1990 | 0.13 | 2.60 | 27.02 | 4.85 | 0.49 | 0.32 |  |  | 35.41 |
| 1991 | 0.00 | 5.00 | 19.83 | 32.67 | 2.75 | 0.19 | 0.17 |  | 60.61 |
| 1992 | 2.74 | 5.23 | 20.80 | 20.87 | 79.60 | 4.17 | 1.61 | 0.22 | 135.24 |
| 1993 | 4.87 | 14.58 | 17.35 | 20.22 | 25.44 | 41.95 | 4.74 | $0.71{ }^{\prime}$ | 129.86 |
| 1994 | 23.78 | 25.85 | 10.36 | 8.21 | 7.68 | 3.49 | 17.53 | $2.61{ }^{\prime \prime}$ | 99.51 |
| 1995 | 6.49 | 35.24 | 12.34 | 2.27 | 3.60 | 2.56 | 2.15 | 7.96 | 72.61 |
| 1996 | 1.41 | 14.43 | 24.00 | 3.65 | 0.79 | 0.25 | 0.80 | 1.30 | 46.63 |
| 1997 | 0.40 | 4.95 | 27.56 | 16.50 | 1.50 | 0.42 |  | 0.75 | 52.08 |
| 1998 | 0.05 | 0.30 | 7.06 | 11.05 | 3.24 | 0.51 | 0.18 | $0.02{ }^{\prime \prime}$ | 22.41 |
| 1999 | 0.25 | 1.92 | 4.84 | 14.58 | 8.42 | 0.75 | 0.19 | $0.10^{\prime \prime}$ | 31.05 |
| 2000 | 3.61 | 3.85 | 3.25 | 2.15 | 2.23 | 0.45 | 0.39 | $0.05{ }^{\prime \prime}$ | 15.98 |
| 2001 | 4.33 | 17.61 | 8.03 | 0.96 | 0.33 | 0.36 | 0.26 | $0.09{ }^{\prime \prime}$ | 31.97 |
| 2002 | 2.30 | 19.11 | 16.50 | 6.49 | 0.83 | 0.31 | 0.47 | $0.01{ }^{\prime \prime}$ | 46.02 |
| 2003 | 2.49 | 29.56 | 30.01 | 13.46 | 1.90 | 0.11 | 0.04 | $0.02{ }^{\prime \prime}$ | 77.59 |
| 2004 | 1.96 | 17.52 | 29.82 | 16.34 | 7.67 | 2.04 | 0.15 | $0.68{ }^{\prime \prime}$ | 76.18 |
| 2005 | 3.33 | 12.93 | 28.75 | 13.06 | 6.51 | 1.55 | 0.06 | 0.16 | 66.35 |
| 2006 | 0.20 | 12.50 | 8.11 | 10.98 | 7.42 | 2.12 | 0.16 | 0.66 | 42.14 |
| 2007 | 1.46 | 3.88 | 28.52 | 8.69 | 5.35 | 2.80 | 0.68 | 0.36 | 51.72 |
| 2008 | 0.45 | 5.96 | 2.95 | 20.72 | 2.70 | 2.02 | 1.66 | 0.71 | 37.17 |
| 2009 | 3.42 | 14.48 | 27.64 | 8.10 | 22.31 | 3.07 | 1.56 | $0.37{ }^{\prime \prime}$ | 80.95 |
| 2010 | 1.22 | 32.60 | 26.50 | 23.68 | 7.56 | 6.32 | 0.81 | 1.54 | 100.22 |

Table A5. North-east Arctic COD. Mean length at age(cm) from Norwegian surveys in January-March 1983-1999 values re-calculated from raw data.

|  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| 1978 | 14.2 | 23.1 | 32.1 | 45.9 | 54.2 | 64.6 | 67.6 | 76.9 |
| 1979 | 12.8 | 22.9 | 33.1 | 40.0 | 52.3 | 64.4 | 74.7 | 83.0 |
| 1980 | 17.6 | 24.8 | 34.2 | 40.5 | 52.5 | 63.5 | 73.6 | 83.6 |
| 1981 | 17.0 | 26.1 | 35.5 | 44.7 | 52.0 | 61.3 | 69.6 | 77.9 |
| 1982 | 14.8 | 25.8 | 37.6 | 46.3 | 54.7 | 63.1 | 70.8 | 82.9 |
| 1983 | 12.8 | 27.6 | 34.8 | 45.9 | 54.5 | 62.7 | 73.1 | 78.6 |
| 1984 | 14.2 | 28.4 | 35.8 | 48.6 | 56.6 | 66.2 | 74.1 | 79.7 |
| 1985 | 16.5 | 23.7 | 40.3 | 48.7 | 61.3 | 71.1 | 81.2 | 85.7 |
| 1986 | 11.9 | 21.6 | 34.4 | 49.9 | 59.8 | 69.4 | 80.3 | 93.8 |
| 1987 | 13.9 | 21.0 | 31.8 | 41.3 | 56.3 | 66.3 | 77.6 | 87.9 |
| 1988 | 15.3 | 23.3 | 29.7 | 38.7 | 47.6 | 56.8 | 71.7 | 79.4 |
| 1989 | 12.5 | 25.4 | 34.7 | 39.9 | 46.8 | 56.2 | 67.0 | 83.3 |
| 1990 | 14.4 | 27.9 | 39.4 | 47.1 | 53.8 | 60.6 | 68.2 | 79.2 |
| 1991 | 13.6 | 27.2 | 41.6 | 51.7 | 59.5 | 67.1 | 72.3 | 77.6 |
| 1992 | 13.2 | 23.9 | 41.3 | 49.9 | 60.2 | 68.4 | 76.1 | 82.8 |
| 1993 | 11.3 | 20.3 | 35.9 | 50.8 | 59.0 | 68.2 | 76.8 | 85.8 |
| 1994 | 12.0 | 18.3 | 30.5 | 44.7 | 55.4 | 64.3 | 73.5 | 82.4 |
| 1995 | 12.7 | 18.7 | 29.9 | 42.0 | 54.1 | 64.1 | 74.8 | 80.6 |
| 1996 | 12.6 | 19.6 | 28.1 | 41.0 | 49.3 | 61.4 | 72.2 | 85.3 |
| 1997 | 11.4 | 18.8 | 28.0 | 40.4 | 49.9 | 59.3 | 69.1 | 80.6 |
| $1998{ }^{1}$ | 10.9 | 17.4 | 28.7 | 40.0 | 50.5 | 58.9 | 67.5 | 76.3 |
| 1999 | 12.1 | 18.8 | 29.0 | 40.6 | 50.6 | 59.9 | 70.3 | 78.0 |
| 2000 | 13.0 | 21.0 | 28.7 | 39.7 | 51.5 | 61.6 | 70.5 | 75.7 |
| 2001 | 12.0 | 22.5 | 33.1 | 41.6 | 52.2 | 63.1 | 71.2 | 79.2 |
| 2002 | 12.2 | 19.9 | 30.1 | 43.6 | 52.2 | 61.7 | 71.6 | 79.1 |
| 2003 | 12.0 | 21.2 | 29.1 | 39.2 | 53.3 | 61.6 | 70.3 | 80.7 |
| 2004 | 11.0 | 18.9 | 32.0 | 40.9 | 52.0 | 61.8 | 69.0 | 79.0 |
| 2005 | 11.5 | 18.6 | 29.3 | 43.0 | 51.1 | 60.3 | 71.1 | 78.4 |
| 2006 | 12.2 | 19.9 | 31.3 | 42.1 | 53.5 | 60.8 | 68.9 | 77.7 |
| 2007 | 13.4 | 21.3 | 30.7 | 42.2 | 52.8 | 62.3 | 70.5 | 77.9 |
| 2008 | 12.5 | 22.3 | 32.5 | 43.7 | 52.4 | 63.6 | 71.6 | 80.8 |
| 2009 | 11.7 | 21.4 | 32.2 | 43.2 | 53.6 | 63.3 | 76.0 | 84.4 |
| 2010 | 11.4 | 19.1 | 31.2 | 42.3 | 52.0 | 61.3 | 70.5 | 80.6 |

Table A6. North-east Arctic COD. Weight (g) at age from Norwegian surveys in January-March Year

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1983 |  | 190 | 372 | 923 | 1597 | 2442 | 3821 | 4758 |
| 1984 | 23 | 219 | 421 | 1155 | 1806 | 2793 | 3777 | 4566 |
| 1985 |  | 171 | 576 | 1003 | 2019 | 3353 | 5015 | 6154 |
| 1986 |  | 119 | 377 | 997 | 1623 | 2926 | 3838 | 7385 |
| $1987{ }^{2}$ | 21 | 65 | 230 | 490 | 1380 | 2300 | 3970 |  |
| 1988 | 24 | 114 | 241 | 492 | 892 | 1635 | 3040 | 4373 |
| 1989 | 16 | 158 | 374 | 604 | 947 | 1535 | 2582 | 4906 |
| 1990 | 26 | 217 | 580 | 1009 | 1435 | 1977 | 2829 | 4435 |
| 1991 | 18 | 196 | 805 | 1364 | 2067 | 2806 | 3557 | 4502 |
| 1992 | 20 | 136 | 619 | 1118 | 1912 | 2792 | 3933 | 5127 |
| 1993 | 9 | 71 | 415 | 1179 | 1743 | 2742 | 3977 | 5758 |
| 1994 | 13 | 55 | 259 | 788 | 1468 | 2233 | 3355 | 4908 |
| 1995 | 16 | 54 | 248 | 654 | 1335 | 2221 | 3483 | 4713 |
| 1996 | 15 | 62 | 210 | 636 | 1063 | 1999 | 3344 | 5514 |
| $1997{ }^{1}$ | 12 | 54 | 213 | 606 | 1112 | 1790 | 2851 | 4761 |
| $1998{ }^{1}$ | 10 | 47 | 231 | 579 | 1145 | 1732 | 2589 | 3930 |
| 1999 | 13 | 55 | 219 | 604 | 1161 | 1865 | 2981 | 3991 |
| 2000 | 17 | 77 | 210 | 559 | 1189 | 1978 | 2989 | 3797 |
| 2001 | 14 | 103 | 338 | 664 | 1257 | 2188 | 3145 | 4463 |
| 2002 | 15 | 68 | 256 | 747 | 1234 | 2024 | 3190 | 4511 |
| 2003 | 14 | 82 | 228 | 569 | 1302 | 1980 | 2975 | 4666 |
| 2004 | 11 | 58 | 294 | 600 | 1167 | 1934 | 2657 | 4025 |
| 2005 | 13 | 57 | 230 | 705 | 1135 | 1817 | 2948 | 4081 |
| 2006 | 15 | 71 | 288 | 682 | 1366 | 1991 | 2959 | 4354 |
| 2007 | 19 | 78 | 253 | 691 | 1302 | 2128 | 3032 | 4327 |
| 2008 | 16 | 94 | 319 | 798 | 1393 | 2412 | 3413 | 5067 |
| 2009 | 13 | 83 | 291 | 724 | 1337 | 2180 | 3775 | 5267 |
| 2010 | 12 | 63 | 300 | 683 | 1246 | 2041 | 3076 | 4765 |
| ${ }^{1}$ Adjusted weights <br> ${ }^{2}$ Estimated weights |  |  |  |  |  |  |  |  |

Table A7. Northeast Arctic COD. Length at age in cm in the Lofoten survey

| Year/age | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | 59.6 | 71.1 | 79.0 | 88.2 | 97.3 | 105.2 | 114.0 |  |
| 1986 | 62.7 | 70.0 | 80.0 | 89.4 | 86.6 |  | 105.8 | 115.0 |
| 1987 | 58.2 | 64.5 | 76.7 | 86.2 | 88.0 |  | 118.5 | 116.0 |
| 1988 | 53.1 | 67.1 | 71.6 | 94.0 | 97.0 | 119.6 |  |  |
| 1989 | 54.0 | 59.0 | 69.8 | 80.8 | 96.6 | 103.0 |  | 125.0 |
| 1990 | 56.9 | 65.1 | 69.2 | 79.5 | 83.7 | 100.1 |  |  |
| 1991 | 59.0 | 67.3 | 74.4 | 81.0 | 91.3 | 99.8 | 85.0 |  |
| 1992 | 66.3 | 68.7 | 78.3 | 83.9 | 89.2 | 92.2 | 101.9 | 127.0 |
| 1993 | 58.3 | 66.1 | 72.8 | 83.6 | 87.4 | 92.7 | 95.4 | 111.2 |
| 1994 | 64.3 | 70.6 | 82.0 | 87.3 | 90.0 | 95.3 | 92.4 | 101.4 |
| 1995 | 61.5 | 69.7 | 77.8 | 84.4 | 92.6 | 96.7 | 100.3 | 99.5 |
| 1996 | 62.2 | 67.1 | 75.9 | 81.0 | 93.6 | 100.9 | 97.4 | 104.1 |
| 1997 | 63.7 | 68.6 | 74.2 | 83.8 | 99.9 | 108.4 |  | 109.0 |
| 1998 | 55.0 | 62.6 | 70.2 | 80.0 | 92.0 | 98.0 | 96.7 | 115.0 |
| 1999 | 52.7 | 67.0 | 69.4 | 78.6 | 85.8 | 100.3 | 102.0 | 125.0 |
| 2000 | 58.4 | 66.5 | 72.6 | 77.0 | 83.9 | 90.6 | 93.7 | 112.4 |
| 2001 | 59.3 | 66.9 | 73.2 | 87.1 | 88.7 | 102.8 | 98.5 | 128.2 |
| 2002 | 58.6 | 66.0 | 73.2 | 80.8 | 88.2 | 101.8 | 91.0 | 101.4 |
| 2003 | 62.3 | 65.0 | 73.2 | 80.9 | 88.9 | 86.4 | 120.0 | 122.0 |
| 2004 | 58.8 | 64.7 | 71.2 | 80.1 | 85.6 | 97.0 | 102.6 | 115.8 |
| 2005 | 56.3 | 65.4 | 72.3 | 76.0 | 85.3 | 95.5 | 110.5 | 117.8 |
| 2006 | 56.2 | 63.7 | 72.6 | 77.5 | 82.9 | 88.3 | 89.2 | 116.3 |
| 2007 | 63.0 | 66.4 | 72.4 | 82.5 | 88.2 | 99.8 | 103.7 | 115.0 |
| 2008 | 63.8 | 69.1 | 73.6 | 80.9 | 90.0 | 94.9 | 94.9 | 96.5 |
| 2009 | 60.5 | 69.3 | 76.5 | 82.7 | 88.7 | 98.8 | 92.9 | 111.6 |
| 2010 | 60.6 | 64.2 | 75.0 | 82.8 | 93.9 | 93.7 | 102.8 | 108.1 |

Table A8. Northeast Arctic COD. Mean weight at age (kg) in the Lofoten survey

| Year | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1985 | 2.00 | 3.42 | 4.61 | 6.67 | 8.89 | 10.73 | 14.29 |  |
| 1986 | 2.22 | 3.22 | 4.74 | 6.40 | 5.80 |  | 10.84 | 13.48 |
| 1987 | 1.44 | 1.94 | 3.61 | 5.40 | 5.64 |  | 13.15 | 12.55 |
| 1988 | 1.46 | 2.82 | 3.39 | 6.63 | 7.27 | 13.64 |  |  |
| 1989 | 1.30 | 1.77 | 2.89 | 4.74 | 8.28 | 9.98 |  | 26.00 |
| 1990 | 1.54 | 2.32 | 2.55 | 3.78 | 4.77 | 8.80 |  |  |
| 1991 | 2.21 | 2.52 | 3.51 | 5.18 | 7.40 | 11.36 | 5.35 |  |
| 1992 | 2.56 | 2.85 | 3.99 | 5.43 | 6.35 | 8.03 | 9.50 | 17.80 |
| 1993 | 1.79 | 2.58 | 3.55 | 5.31 | 6.21 | 7.69 | 9.28 | 14.71 |
| 1994 | 2.31 | 3.27 | 5.06 | 6.39 | 6.64 | 7.92 | 7.73 | 10.10 |
| 1995 | 2.20 | 3.24 | 4.83 | 5.98 | 7.80 | 10.03 | 10.39 | 10.68 |
| 1996 | 2.22 | 2.75 | 4.11 | 5.63 | 7.92 | 10.53 | 10.58 | 12.08 |
| 1997 | 2.42 | 2.92 | 3.86 | 5.71 | 9.65 | 13.41 |  | 12.67 |
| 1998 | 1.88 | 2.09 | 2.98 | 4.85 | 7.92 | 9.91 | 11.05 | 18.34 |
| 1999 | 1.51 | 2.80 | 2.96 | 4.22 | 5.92 | 9.33 | 9.17 | 16.00 |
| 2000 | 1.71 | 2.50 | 3.16 | 3.85 | 5.32 | 7.07 | 7.62 | 12.84 |
| 2001 | 1.90 | 2.72 | 3.49 | 6.23 | 6.82 | 10.95 | 10.29 | 28.58 |
| 2002 | 1.87 | 2.57 | 3.52 | 4.71 | 6.18 | 10.56 | 8.70 | 10.48 |
| 2003 | 2.30 | 2.34 | 3.48 | 4.59 | 5.89 | 8.07 | 24.50 | 27.70 |
| 2004 | 1.74 | 2.30 | 3.02 | 4.50 | 5.77 | 7.81 | 9.95 | 13.25 |
| 2005 | 1.56 | 2.40 | 3.20 | 3.71 | 5.79 | 8.52 | 16.27 | 18.63 |
| 2006 | 1.54 | 2.35 | 3.44 | 4.19 | 5.43 | 6.57 | 6.19 | 18.15 |
| 2007 | 2.34 | 2.67 | 3.53 | 5.30 | 6.70 | 9.95 | 11.24 | 16.62 |
| 2008 | 2.21 | 2.97 | 3.63 | 4.88 | 6.74 | 8.18 | 7.70 | 9.07 |
| 2009 | 2.04 | 2.98 | 4.1 | 5.19 | 6.56 | 9.38 | 8.58 | 15.67 |
| 2010 | 1.91 | 2.28 | 3.60 | 4.70 | 7.03 | 7.11 | 9.09 | 12.50 |

Table A9 North-east Arctic COD. Results from the Russian trawl-acoustic survey in the Barents Sea and adjacent wates in the autumn. Stock number in millions.

| Year |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | Total |
| $1985{ }^{1}$ | 77 | 569 | 400 | 568 | 244 | 51 | 20 | 8 | 1 | 3 | 1941 |
| $1986{ }^{1}$ | 25 | 129 | 899 | 612 | 238 | 69 | 20 | 3 | 2 | 1 | 1998 |
| $1987{ }^{2}$ | 2 | 58 | 103 | 855 | 198 | 82 | 19 | 4 | 1 | 1 | 1323 |
| $1988{ }^{2}$ | 3 | 23 | 96 | 100 | 305 | 54 | 16 | 3 | 1 | 1 | 602 |
| $1989{ }^{1}$ | 1 | 3 | 17 | 45 | 57 | 91 | 75 | 25 | 13 | 5 | 332 |
| $1990{ }^{1}$ | 36 | 27 | 8 | 27 | 62 | 74 | 91 | 39 | 10 | 3 | 377 |
| $1991{ }^{1}$ | 63 | 65 | 96 | 45 | 50 | 54 | 66 | 49 | 5 | 1 | 494 |
| $1992{ }^{1}$ | 133 | 399 | 380 | 121 | 56 | 58 | 33 | 29 | 11 | 2 | 1222 |
| $1993{ }^{1}$ | 20 | 44 | 220 | 234 | 164 | 51 | 19 | 13 | 8 | $10^{\prime \prime}$ | 783 |
| $1994{ }^{1}$ | 105 | 38 | 147 | 275 | 303 | 314 | 100 | 35 | 10 | 8 | 1335 |
| $1995{ }^{1}$ | 242 | 42 | 111 | 219 | 229 | 97 | 21 | 6 | 2 | 2 | 971 |
| $1996{ }^{1,3,5}$ | 424 | 275 | 189 | 316 | 449 | 314 | 126 | 27 | 3 | 4 | 2127 |
| $1997{ }^{4,5}$ | 72 | 160 | 263 | 198 | 112 | 57 | 27 | 9 | 1 | 1 | 900 |
| $1998{ }^{1}$ | 26 | 86 | 279 | 186 | 57 | 23 | 10 | 4 | 1 | 0 | 672 |
| $1999{ }^{1}$ | 19 | 79 | 166 | 260 | 98 | 20 | 8 | 5 | 2 | $1{ }^{\prime \prime}$ | 658 |
| $2000{ }^{1, \text { rev }}$ | 24 | 82 | 191 | 159 | 127 | 48 | 6 | 3 | 1 | 1 | 642 |
| $2001{ }^{1}$ | 38 | 59 | 148 | 204 | 120 | 70 | 14 | 2 | 1 |  | 656 |
| $2002{ }^{1,5,6}$ | 83 | 2 | 106 | 85 | 140 | 151 | 67 | 30 | 7 | 1 | 672 |
| 2003 | 69 | 36 | 25 | 218 | 142 | 167 | 163 | 60 | 23 | 4 | 908 |
| 2004 | 375 | 35 | 170 | 85 | 345 | 194 | 229 | 167 | 49 | 19 | 1669 |
| 2005 | 112 | 48 | 65 | 154 | 70 | 214 | 68 | 47 | 17 | 8 | 803 |
| $2006{ }^{7}$ | 12 | 20 | 39 | 49 | 78 | 32 | 64 | 23 | 13 | 8 | 341 |
| 2007 | 13 | 35 | 165 | 372 | 208 | 189 | 74 | 113 | 32 | 20 | 1221 |
| ${ }^{1}$ October-December |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{2}$ September-October |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{3}$ Area llb not covered |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{4}$ Areas lla, llb covered in October-December, part of Area I covered in February-March 1998 |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{5}$ Adjusted for incomplete area coverage |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{6}$ Area lla not covered |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{7}$ Area I not fully covered |  |  |  |  |  |  |  |  |  |  |  |

Table A10. North-East Arctic COD. Abundance indices (millions) from the Russian bottom trawl survey in the Barents Sea


Table A11 North-East Arctic COD. Length at age (cm) from Russian surveys in Novem-ber-December

| Year | Age |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |  |
| 1984 | 15.7 | 22.3 | 30.7 | 44.3 | 51.7 | 63.6 | 73.4 | 82.5 | 88.4 | 97.0 |  |
| 1985 | 15.0 | 21.1 | 30.6 | 43.2 | 53.7 | 61.2 | 72.8 | 83.0 | 92.8 | 101.3 |  |
| 1986 | 15.2 | 19.7 | 28.3 | 39.0 | 51.8 | 62.2 | 70.9 | 83.0 | 91.3 | 104.0 |  |
| 1987 | - | 19.2 | 27.9 | 33.4 | 41.4 | 59.1 | 69.2 | 80.1 | 95.7 | 102.6 |  |
| 1988 | 11.3 | 21.3 | 28.7 | 36.2 | 43.9 | 53.3 | 65.3 | 79.5 | 85.0 | - |  |
| 1989 | - | 20.8 | 28.8 | 34.8 | 46.0 | 53.9 | 61.8 | 69.7 | 69.8 | 78.7 |  |
| 1990 | 16.0 | 24.0 | 30.4 | 46.5 | 54.9 | 62.5 | 72.8 | 77.6 | 87.8 | 102.6 |  |
| 1991 | 11.5 | 22.4 | 30.6 | 43.0 | 55.9 | 64.6 | 78.6 | 78.5 | 87.9 | 101.8 |  |
| 1992 | 11.3 | 21.3 | 31.9 | 50.1 | 59.8 | 69.1 | 73.9 | 84.0 | 90.8 | 97.5 |  |
| 1993 | 12.1 | 17.4 | 29.1 | 43.4 | 52.7 | 64.3 | 70.6 | 81.2 | 89.1 | 91.8 |  |
| 1994 | 12.2 | 20.3 | 26.3 | 33.7 | 47.4 | 58.7 | 71.1 | 80.8 | 90.1 | 96.1 |  |
| 1995 | 11.6 | 19.8 | 27.6 | 33.8 | 45.2 | 60.5 | 70.5 | 83.5 | 92.9 | 99.1 |  |
| 1996 | 10.2 | 20.0 | 28.1 | 36.7 | 48.7 | 58.9 | 70.5 | 80.0 | 93.6 | 102.7 |  |
| 1997 | 9.6 | 18.5 | 28.8 | 38.2 | 50.8 | 62.0 | 70.7 | 80.1 | 88.9 | 103.5 |  |
| 1998 | 11.4 | 19.0 | 28.0 | 36.4 | 50.5 | 61.0 | 71.6 | 80.3 | 91.1 | 102.5 |  |
| 1999 | 11.7 | 19.7 | 27.9 | 35.3 | 51.6 | 60.6 | 71.9 | 78.9 | 86.8 | 94.3 |  |
| 2000 | 10.7 | 20.8 | 30.1 | 34.7 | 49.8 | 61.1 | 70.6 | 82.0 | 88.3 | 85.7 |  |
| 2001 | 10.6 | 19.4 | 29.8 | 37.3 | 50.4 | 61.9 |  | 81.4 | 91.0 | 98.7 |  |
| 2002 | 10.7 | 19.2 | 29.9 | 38.2 | 52.5 | 60.4 |  | 82.2 | 91.3 | 97.2 |  |
| 2003 | 9.8 | 18.9 | 28.3 | 34.9 | 49.2 | 62.2 | 71.0 | 81.5 | 92.3 | 100.9 |  |
| 2004 | 9.8 | 19.6 | 29.3 | 38.4 | 49.1 | 60.0 | 70.5 | 80.0 | 91.0 | 98.0 |  |
| 2005 | 11.2 | 19.4 | 29.7 | 38.5 | 48.7 | 59.3 | 69.3 | 79.2 | 87.7 | 96.1 |  |
| 2006 | 13.0 | 21.9 | 31.6 | 42.7 | 53.2 | 60.1 | 70.2 | 79.1 | 88.3 | 95.2 |  |
| 2007 | 10.7 | 21.5 | 30.8 | 42.2 | 53.6 | 63.7 | 71.0 | 79.6 | 87.3 | 95.9 |  |
| 2008 | 10.2 | 20.0 | 30.3 | 40.2 | 53.7 | 64.5 | 74.6 | 82.7 | 89.5 | 98.2 |  |
| 2009 | 12.9 | 19.3 | 29.5 | 38.4 | 50.7 | 61.5 | 70.7 | 81.7 | 89.9 | 94.7 |  |

Table A12 North-East Arctic COD. Weight (g) at age from Russian surveys in Novem-ber-December.

| Year | Age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| 1984 | 26 | 90 | 250 | 746 | 1,187 | 2,234 | 3,422 | 5,027 | 6,479 | 9,503 | - |
| 1985 | 26 | 80 | 245 | 762 | 1,296 | 1,924 | 3,346 | 5,094 | 7,360 | 6,833 | 11,167 |
| 1986 | 25 | 63 | 191 | 506 | 1,117 | 1,940 | 2,949 | 4,942 | 7,406 | 9,300 | - |
| 1987 | - | 54 | 182 | 316 | 672 | 1,691 | 2,688 | 3,959 | 8,353 | 10,583 | 13,107 |
| 1988 | 15 | 78 | 223 | 435 | 789 | 1,373 | 2,609 | 4,465 | 5,816 | - | - |
| 1989 | - | 73 | 216 | 401 | 928 | 1,427 | 2,200 | 3,133 | 4,649 | 6,801 | 8,956 |
| 1990 | 28 | 106 | 230 | 908 | 1,418 | 2,092 | 2,897 | 4,131 | 6,359 | 10,078 | 13,540 |
| 1991 | 26 | 93 | 260 | 743 | 1,629 | 2,623 | 3,816 | 4,975 | 7,198 | 11,165 | 15,353 |
| 1992 | 10 | 76 | 273 | 1,165 | 1,895 | 2,971 | 4,377 | 5,596 | 7,319 | 9,452 | 12,414 |
| 1993 | 11 | 46 | 211 | 717 | 1,280 | 2,293 | 3,509 | 4,902 | 6,621 | 7,339 | 8,494 |
| 1994 | 12 | 69 | 153 | 316 | 919 | 1,670 | 2,884 | 4,505 | 6,520 | 8,207 | 9,812 |
| 1995 | 11 | 61 | 180 | 337 | 861 | 1,987 | 3,298 | 5,427 | 7,614 | 9,787 | 10,757 |
| 1996 | 7 | 64 | 191 | 436 | 1,035 | 1,834 | 3,329 | 5,001 | 8,203 | 10,898 | 11,358 |
| 1997 | 6 | 48 | 203 | 487 | 1,176 | 2,142 | 3,220 | 4,805 | 6,925 | 10,823 | 12,426 |
| 1998 | 11 | 55 | 187 | 435 | 1,186 | 2,050 | 3,096 | 4,759 | 7,044 | 11,207 | 12,593 |
| 1999 | 10 | 58 | 177 | 371 | 1,214 | 1,925 | 3,064 | 4,378 | 6,128 | 7,843 | 11,543 |
| 2000 | 8 | 74 | 232 | 379 | 1,101 | 2,128 | 3,341 | 5,054 | 6,560 | 8,497 | 12,353 |
| 2001 | 9 | 58 | 221 | 459 | 1,125 | 2,078 | 3,329 | 4,950 | 7,270 | 9,541 | 11,672 |
| 2002 | 8 | 65 | 232 | 505 | 1,299 | 1,964 | 3,271 | 5,325 | 7,249 | 9,195 | 11,389 |
| 2003 | 6 | 49 | 205 | 492 | 972 | 1,993 | 2,953 | 4,393 | 6,638 | 9,319 | 11,085 |
| 2004 | 6 | 55 | 231 | 543 | 1,079 | 1,798 | 2,977 | 4,110 | 5,822 | 8,061 | 12,442 |
| 2005 | 10 | 59 | 223 | 521 | 1,034 | 1,910 | 3,036 | 4,619 | 6,580 | 9,106 | 12,006 |
| 2006 | 13 | 72 | 270 | 707 | 1,332 | 1,953 | 2,969 | 4,340 | 6,410 | 8,622 | 12,436 |
| 2007 | 10 | 96 | 252 | 669 | 1,344 | 2,277 | 3,140 | 4,691 | 6,178 | 8,567 | 10,014 |
| 2008 | 7 | 58 | 228 | 558 | 1,332 | 2,305 | 3,527 | 5,001 | 6,519 | 8,848 | 10,339 |
| 2009 | 15 | 54 | 214 | 495 | 1,116 | 2,024 | 3,090 | 4,876 | 6,592 | 8,087 | 10,262 |

Table A13. North-East Arctic COD. Sum of acoustic abundance estimates (millions) in the Joint winter Barents Sea survey (Table A2) and the Norwegian Lo foten acoustic survey (Table A4)

|  | Age |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| 1985 | 69.1 | 446.3 | 153.0 | 141.6 | 20.4 | 15.1 | 15.7 | 3.3 | 1.3 | 1.0 | 0.5 | 0.0 |
| 1986 | 353.6 | 243.9 | 499.6 | 134.3 | 68.4 | 11.6 | 7.7 | 3.1 | 0.3 | 0.0 | 0.4 | 0.1 |
| 1987 | 1.6 | 34.1 | 62.8 | 204.9 | 50.2 | 17.4 | 1.4 | 3.0 | 0.7 | 0.0 | 0.0 | 0.0 |
| 1988 | 2.0 | 26.3 | 50.4 | 35.5 | 57.8 | 10.9 | 4.0 | 0.3 | 0.0 | 0.1 | 0.0 | 0.0 |
| 1989 | 7.5 | 8.0 | 17.0 | 34.4 | 21.4 | 67.0 | 16.6 | 3.2 | 0.5 | 0.2 | 0.0 | 0.1 |
| 1990 | 81.1 | 24.9 | 14.8 | 20.6 | 26.2 | 26.9 | 66.8 | 7.3 | 0.6 | 0.3 | 0.0 | 0.0 |
| 1991 | 181.0 | 219.5 | 50.2 | 34.6 | 29.3 | 33.9 | 36.7 | 50.0 | 3.7 | 0.2 | 0.2 | 0.0 |
| 1992 | 241.4 | 562.1 | 176.5 | 65.8 | 21.5 | 18.4 | 28.4 | 25.4 | 82.4 | 4.3 | 1.7 | 0.2 |
| 1993 | 1074.0 | 494.7 | 357.2 | 191.1 | 113.1 | 35.4 | 25.5 | 25.2 | 27.7 | 44.2 | 4.9 | 0.8 |
| 1994 | 858.3 | 577.2 | 349.8 | 404.5 | 217.5 | 89.5 | 22.5 | 11.9 | 9.4 | 3.9 | 18.0 | 2.7 |
| 1995 | 2619.2 | 292.9 | 166.2 | 159.8 | 216.6 | 104.0 | 29.0 | 4.4 | 4.3 | 3.0 | 2.6 | 8.1 |
| 1996 | 2396.0 | 339.8 | 92.9 | 70.5 | 87.2 | 89.1 | 44.6 | 6.5 | 1.1 | 0.4 | 0.9 | 1.4 |
| 1997 | 1623.5 | 430.5 | 188.3 | 51.7 | 49.7 | 42.2 | 49.9 | 20.5 | 2.2 | 0.5 | 0.0 | 0.8 |
| 1998 | 3401.3 | 632.9 | 427.7 | 182.6 | 42.4 | 33.8 | 34.0 | 24.7 | 4.9 | 0.7 | 0.2 | 0.1 |
| 1999 | 358.3 | 304.3 | 150.0 | 96.4 | 45.4 | 12.2 | 11.2 | 18.7 | 9.2 | 1.0 | 0.2 | 0.2 |
| 2000 | 154.1 | 221.4 | 245.2 | 158.9 | 145.7 | 49.3 | 12.9 | 6.9 | 5.2 | 1.2 | 0.6 | 0.2 |
| 2001 | 629.9 | 63.9 | 138.2 | 171.6 | 81.6 | 57.3 | 19.8 | 2.4 | 0.8 | 0.6 | 0.3 | 0.1 |
| 2002 | 18.2 | 215.5 | 69.3 | 112.2 | 104.3 | 66.1 | 34.5 | 9.5 | 1.2 | 0.5 | 0.6 | 0.0 |
| 2003 | 1693.9 | 61.5 | 303.4 | 114.4 | 131.5 | 144.5 | 64.3 | 21.2 | 3.8 | 0.5 | 0.1 | 0.1 |
| 2004 | 157.7 | 105.2 | 33.6 | 92.8 | 32.7 | 45.1 | 46.8 | 22.2 | 8.8 | 2.2 | 0.2 | 0.7 |
| 2005 | 465.3 | 119.6 | 123.9 | 33.7 | 66.1 | 29.9 | 43.2 | 17.2 | 7.5 | 1.8 | 0.1 | 0.2 |
| 2006 | 544.6 | 216.6 | 79.8 | 59.1 | 15.7 | 38.1 | 16.9 | 15.5 | 8.8 | 2.4 | 0.3 | 0.8 |
| 2007 | 125.0 | 61.7 | 80.3 | 37.1 | 31.8 | 13.0 | 42.7 | 13.8 | 7.5 | 3.3 | 0.8 | 0.4 |
| 2008 | 68.8 | 97.6 | 210.2 | 306.1 | 141.0 | 75.4 | 24.6 | 32.9 | 5.8 | 2.8 | 1.7 | 0.8 |
| 2009 | 321.5 | 30.6 | 182.6 | 178.3 | 140.5 | 49.5 | 40.1 | 13.3 | 26.0 | 3.7 | 1.7 | 0.4 |
| 2010 | 485.4 | 59.4 | 34.7 | 121.9 | 175.9 | 194.9 | 70.9 | 37.5 | 11.1 | 8.8 | 1.7 | 1.7 |

Table A14. Northeast Arctic Cod. Swept area estimates (millions) from the Joint NorwegianRussian ecosystem survey in August-September (taken from WD 20)

| year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | $13+$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2004 | 540.45 | 332.75 | 329.74 | 147.72 | 421.53 | 150.21 | 79.76 | 40.21 | 10.09 | 2.21 | 0.50 | 0.13 | 0.07 | 0.13 |
| 2005 | 182.17 | 458.52 | 143.16 | 241.68 | 95.92 | 159.91 | 35.54 | 16.24 | 5.82 | 1.01 | 0.47 | 0.17 | 0.00 | 0.06 |
| 2006 | 274.55 | 479.02 | 509.66 | 186.11 | 205.71 | 59.96 | 70.25 | 17.93 | 8.21 | 2.58 | 0.65 | 0.25 | 0.04 | 0.00 |
| 2007 | 97.79 | 334.23 | 506.27 | 587.08 | 159.42 | 79.26 | 24.68 | 27.15 | 6.05 | 2.18 | 0.94 | 0.15 | 0.21 | 0.03 |
| 2008 | 493.55 | 131.01 | 372.92 | 654.33 | 486.23 | 133.06 | 51.79 | 12.93 | 17.57 | 3.30 | 0.85 | 0.23 | 0.20 | 0.19 |
| 2009 | 922.67 | 580.99 | 91.20 | 202.50 | 286.83 | 295.25 | 103.41 | 32.45 | 12.89 | 7.39 | 2.64 | 0.83 | 0.28 | 0.22 |

### 4.1 Status of the Fisheries

### 4.1.1 Historical development of the fisheries

Haddock is mainly fished by trawl as by-catch in the fishery for cod. Also a directed trawl fishery for haddock is conducted and the proportion of total catches taken by this fishery varies between years. On average approximately $33 \%$ of the catch is with conventional gears, mostly longline, which in the past was used almost exclusively by Norway. Some of the longline catch are from a directed fishery, which is restricted by national quotas. In the Norwegian management the quotas are set separately for trawl and other gears. The fishery is also regulated by a minimum landing size, a minimum mesh size in trawls and Danish seine, a maximum by-catch of undersized fish, closure of areas with high density/catches of juveniles and other seasonal and area restrictions.
The exploitation rate of haddock has been variable. The highest fishing mortalities for haddock have occurred at low to intermediate stock levels and historically show little relationship with the exploitation rate of cod, in spite of haddock being primarily caught as by-catch in the cod fishery. How ever, the more restrictive quota regulations introduced around 1990 have resulted in a more similar pattern in the exploitation rate.

### 4.1.2 Landings prior to 2010 (Tables 4.1-4.3, Figure 4.1A)

The official landings (those reported to ICES and contained in the Statlant statistics) for 2008 amount to $155,604 \mathrm{t}$, and the provisional official landings for 2009 are 200,512 t .
In recent years, estimates of unreported catches (IUU catches) of haddock have been added to reported landings for the years 2002 and onwards. In 2007 to 2009 two estimates of IUU catches were available, one Norwegian and one Russian). For those years, the Working Group decided to present both estimates but to base the final assessment on the Norwegian IUU estimates. In 2009, however, a joint NorwegianRussian Analysis Group under the Mixed Norwegian-Russian Fisheries Commission provided joint estimates of IUU catches. Based on these, the AFWG decided to set the IUU estimate for haddock in 2009 to 0 (WD 13). More details on this issue are given in Sections 0.4 and 3.1.3. Before 2002 the W orking Group has no information about IUU catches on haddock, but the WG consider the IUU fisheries prior to 2002 to be low.
In 2006 it was decided to include reported Norw egian landings of haddock from the Norw egian statistical areas 06 and 07 (ICES 2006) (i.e., between $62^{\circ} \mathrm{N}$ and Lofoten) not previously included in the total landings of NEA haddock used as input for this stock assessment (Tables 4.1-4.3). This practice is continued.

### 4.1.3 Catch advice and landings for 2009 and 2010

ACFM recommended to set a TAC lower than 194,000 $t$ for 2009 and the agreed TAC for 2009 was $194,000 t$, applying the agreed harvest control rule. The provisional reported catch in 2009 is $200,512 \mathrm{t}$. In 2006 and 2007 the assessment of haddock was rejected by ACFM and the advices was in both years to set a TAC lower than $130,000 \mathrm{t}$ based on the increase of SSB in 2001-2004 being associated with this catch level. In 2008-2009 the assessment of haddock was accepted on the basis of improvement in
diagnostics and a clearer explanation of the IUU calculation, and the advice was given according to the agreed 1-year harvest control rule (see Section 4.7.2). For 2010, the mixed Norwegian-Russian Fisheries Commission agreed on a TAC of $243,000 \mathrm{t}$, which corresponds to the agreed 1-year harvest control rule (see Section 4.7.2) according to the assessment. The assessment shows a decreasing trend in F from 2007 to 2009. The F in 2009 is thus considered to be a better estimate for $F$ in the present year (2010) than using a three year average. In predictions for 2010-2012, a three year average scaled to F-status quo (Fsq) was used for the distribution of fishing mortality at age (fishing pattern). A Fsq predicts the catch for 2010 to be $269,000 \mathrm{t}$, which is higher than the TAC $(243,000 \mathrm{t})$. The high 2010 catch corresponding to Fsq should not be interpreted as an estimate of a TAC overshoot in 2010.

### 4.2 Status of Research

### 4.2.1 Survey results (Tables B1-B4, 4.9-4.11)

The overall picture seen in the surveys is summarized as follows: the last poor year class is 1997 and the following six year classes all appear to be at or above average abundance. These are followed by three year classes 2004-2006, which all seem to rank among the $6-7$ most abundant year classes in the VPA time series. According to the 0 -group survey, the 2009 year class seems to be stronger than the two preceding year classes.

## Joint Barents Sea winter survey (bottom trawl and acoustics)

The preliminary swept area estimates and acoustic estimates from the Joint winter survey on demersal fish in the Barents Sea in winter 2010 are given in Aglen et al. (WD 15).

Before 2000 this survey was made without participation from Russian vessels, while in 2001-2005 Russian vessels covered important parts of the Russian zone. In 20062007 only Norwegian vessels carried out the survey again and permit to cover the Russian EEZ was not given in 2007, which meant that the 2007 indices had to be adjusted to take into account the incomplete coverage. These adjustments are described in detail in the 2007 report. How ever, in 2008, 2009, and 2010 permit to enter the Russian zone was again given and the survey was conducted according to the standard area coverage. The survey indices and areas covered are given in Tables B1 and B3.
High indices, caused by the period of good recruitment around 1990, can be tracked from year to year in both series and the 1990 year class appears as the strongest for age groups 3-8 until the 2004-2006 year classes arrive. In the 2010 bottom trawl survey, the 2005 and 2006 year classes show an abundance well above that of the 1990 year class at the same age, while for the 2004 year class ( 6 years in 2010) the index is somewhat lower than for the 1990 year class. In the acoustic survey, the indices of the 2004-2005 year classes are higher than for the 1990 year class at age groups $5-6$, while the 2006 year class (age 4) is a bit smaller.

## Russian bottom trawl and acoustic survey

Russia provided indices from the 2009 Barents Sea trawl and acoustic survey (Tables B2 and B4), which was carried out in October-December. The Russian survey shows similar main trends as the Norwegian survey.

From 1995 onwards there has been a substantial change in the method for calculating acoustic indices. The acoustic survey is therefore presented in 2 tables, Table B4a and B4b, for the old and the new method of calculating indices, respectively.
Also in the Russian bottom trawl and acoustic survey the coverage of REZ in 2006 was reduced compared to previous years, and the survey indices for 2006 were adjusted similar to that of the indices from the joint Barents Sea winter survey. See report from 2007 for details. From 2007 onwards, the survey area covered was again the standard coverage.

## International 0-group survey

Estimates of the abundance of 0-group haddock from the International 0-group survey are presented in Tables 1.1-1.2. The four tables show slightly different pictures, but all indicate that the 2002-2006 year classes are very strong, whereas 2007-2008 year classes are below average. The 2009 year class is again higher, at the level of the 2006 year class.

### 4.2.2 Weight-at-age (Tables B5, B6)

Length- and weight-at-age from the surveys are given in Tables B5 and B6, respectively. Weights-at-age in the Norwegian survey has decreased for the oldest ages compared to last year, whereas the Russian survey shows a decrease for all ages.

### 4.3 Data Used in the Assessment

### 4.3.1 Estimates of unreported catches (Tables 4.1-4.3)

We continue to include the estimates of IUU catches as in previous years (see Section 0.4 and Section 4.1.2), but the IUU estimate is zero for 2009.

### 4.3.2 Catch-at-age (Table 4.4)

Age and length compositions of the landings in 2009 were available from Norway and Russia in Subarea I and Subarea II, and from Norway, Russia, and Germany in Division IIa and Division IIb. The biological sampling of NEA haddock catches is considered to be fairly good. However, the present sampling is believed to be less precise because of the termination of a Norwegian sampling program in Q3 2009 (WD 6). Estimated catch-at-age obtained from Inter catch is listed in Table 4.4.

### 4.3.3 Weight-at-age (Tables 4.5-4.6, Table B.6)

The mean weight-at-age in the catch (Table 4.5) was obtained from Intercatch as a weighted average of the weight at age in the catch for Norway, Russia and Germany. The weights-at-age in the catch in 2009 have increased slightly for younger age groups compared to 2008.

Stock weights (Table 4.6) used from 1985 to 2009 are averages of values derived from Russian surveys in autumn (mostly October-December) and Norwegian surveys in January-March the following year (Table B6). These averages are assumed to give representative values for the beginning of the year. In 2006 the Working group decided to model the stock weight-at-age data in order to remove some of the sampling variability in the estimates. The weight at age is modelled as follows: Mean length at age is modelled using a von Bertalanffy model with $\mathrm{L}_{\infty}$ and $\mathrm{T}_{0}$ parameters estimated over the whole time series and a separate K parameter for each year class. Weight at age is estimated from a length-weight relationship using the smoothed (modelled)
length at age. Estimates were produced separately for the Russian autumn survey and the joint winter survey and were later combined as plain average.

### 4.3.4 Natural mortality (Table 4.7)

Natural mortality used in the assessment was $0.2+$ mortality from predation by cod (see Section 4.4.2). The proportion of F and M before spawning was set to zero. For the period from 1984 to 2009 actual data from predation for cod have been used (see table below) while for the previous years $(1950-1983)$ the average natural mortality for 1984-2009 was used (age groups 1-6).

### 4.3.5 Maturity-at-age (Table 4.8)

In 2006 the Working Group revised the estimates of maturity at age. For the years 1980 onwards the series consists of predicted values using a logistic link function with age and length as explanatory variables from the joint winter survey combined with predicted proportions from the Russian autumn survey:
Mat $=\frac{1}{1+e^{(-a *(a g e-a g e 50 \%)}}$
The new series is based on the data from the Russian autumn survey and the joint winter survey. For the period 1950-1979 an average of both data series is used.

The estimates of maturity-at-age are shown in Table 4.8. The proportions mature at age are presently lower than historic averages.

### 4.3.6 Changes in data from last year (Tables 4.1-4.3)

As stock weights are modelled (See Section 4.3.3) the values of this parameter have been changed slightly both in 1950-1984 for which average values are used and in 1985-2009. The same approach has been used in consumption of NEA haddock by NEA cod estimates and in maturity at age.

### 4.3.7 New data sources

The bottom trawl estimates from the joint ecosystem survey in August-September, starting in 2004. This survey covers the entire distribution area of haddock. The new index values (Table 4.9A) for period 2004-2009 become available for AFWG this year as for the cod (WD 20). This time series have been tested as new tuning fleet in XSA (Fleet 007). The single fleet estimates of survivors of all ages, based on this index were slightly lower compare to other single fleet runs (Figure 4.8). The results of XSA run including all fleets (as in Final XSA + Fleet 007) were very close to final XSA ones. Analysis of XSA diagnostic from this run demonstrate that fleet 007 has reasonably good quality comparable to the other time series (fleet 01,02 and 04 ). This index will be considered on the upcoming benchmark for use as a tuning series.

### 4.4 Assessment Using VPA

The assessment method was also this year XSA.

### 4.4.1 Data for tuning (Table 4.9)

The following surveys series are included in the data for tuning:

| Name | Place | Season | Age | Year | prior <br> weight |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Russian bottom trawl | Barents Sea | Autumn | $1-7$ | $1983-2009$ | 1 |
| Norwegian bottom trawl | Barents Sea | Winter | $1-8$ | $1982-2010$ | 1 |
| Norwegian acoustic | Barents Sea | Winter | $1-7$ | $1980-2010$ | 1 |

The indices for the Russian BT survey in the 1990 were not used for tuning the XSA. Since the 2004 WG meeting the survey data before 1990 have not been used in the XSA run. This decision was based on the analysis of survey residuals and changes in survey methodology (see the 2004 report).

### 4.4.2 VPA and tuning (Table 4.9)

The Extended Survivors Analysis (XSA) was used to tune the VPA to the available index series (Table 4.9). As last years, FLR was used for the assessment of haddock (see 2008-2009 reports), and thus all results concerning XSA is obtained using FLR. The settings used by the AFWG in 2009 were not changed:

The tuning window is set to 20 years
The F shrinkage was given a weight corresponding to $\mathrm{SE}=0.5$
The estimated consumption of NEA haddock by NEA cod is incorporated into the XSA analysis by first constructing a catch number-at-age matrix, adding the numbers of haddock eaten by cod to the catches for the years where such data are available (1984-2009). The consumption of NEA haddock by NEA cod is given below:

|  | Consumption of Haddock by NEA Cod (millions ) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 |
| 1984 | 980.7 | 14.7 | 0.1 | 0.0 | 0.0 | 0.0 |
| 1985 | 1206.2 | 5.2 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1986 | 566.4 | 245.0 | 168.1 | 0.0 | 0.0 | 0.0 |
| 1987 | 768.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1988 | 17.2 | 0.5 | 9.1 | 0.0 | 0.2 | 0.0 |
| 1989 | 230.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1990 | 144.0 | 37.9 | 3.7 | 0.0 | 0.0 | 0.0 |
| 1991 | 457.8 | 14.2 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1992 | 2112.4 | 151.2 | 1.1 | 0.0 | 0.0 | 0.0 |
| 1993 | 1376.6 | 165.7 | 36.8 | 3.4 | 2.9 | 0.0 |
| 1994 | 1412.8 | 80.6 | 25.0 | 7.7 | 0.9 | 0.0 |
| 1995 | 2900.8 | 163.7 | 12.0 | 29.8 | 29.9 | 0.3 |
| 1996 | 1594.1 | 161.4 | 40.2 | 5.5 | 2.6 | 3.4 |
| 1997 | 906.5 | 35.5 | 25.5 | 1.7 | 0.8 | 0.5 |
| 1998 | 1534.8 | 28.2 | 2.0 | 2.9 | 0.5 | 0.0 |
| 1999 | 898.2 | 23.4 | 0.3 | 0.0 | 0.0 | 0.0 |
| 2000 | 1216.4 | 65.0 | 2.1 | 1.1 | 0.2 | 0.1 |
| 2001 | 553.0 | 52.6 | 4.9 | 0.1 | 0.0 | 0.0 |
| 2002 | 2377.1 | 229.1 | 38.0 | 2.4 | 0.4 | 0.2 |
| 2003 | 3616.1 | 219.5 | 38.8 | 12.3 | 1.2 | 0.0 |
| 2004 | 2299.6 | 299.2 | 43.4 | 8.9 | 2.5 | 0.0 |
| 2005 | 5856.7 | 265.4 | 67.3 | 11.9 | 3.5 | 1.2 |
| 2006 | 8012.5 | 335.8 | 3.3 | 4.4 | 1.2 | 0.5 |
| 2007 | 8917.0 | 561.8 | 22.0 | 2.4 | 2.7 | 0.3 |
| 2008 | 1118.1 | 925.7 | 175.5 | 23.5 | 11.6 | 3.1 |
| 2009 | 1295.6 | 201.2 | 191.5 | 39.1 | 17.1 | 1.7 |

The fishing mortality estimated by the XSA was split into the mortality caused by the fishing fleet ( F ) and the mortality caused by the cod's predation (M2) according to the ratio of fleet catch and predation "catch". The new natural mortality data set were then prepared by adding 0.2 (M1) to the predation mortality. This new M matrix (Table 4.7) was used in the final XSA.

The proportion of M and F before spawning was set to 0 .

### 4.4.3 Recruitment indices (Table 4.10, Table 4.11, Figure 4.1 C)

The RCT3 program has been used to estimate the recruiting year-classes 2006-2008 with survey data for ages $0-3$ as input data (Russian autumn survey and joint winter survey). Input data and results are shown in Table 4.10 and 4.11, respectively. Similar to XSA tuning, data points from the 1990 Russian BT were removed from recruitment estimation.

The numbers marked with * are XSA estimates, and the rest are RCT results (Table 4.11). The recruitment time series is shown in Table 4.18 and Figure 4.1C.

| $\mathbf{N}$ | Year of assessment |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Year Class | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| 2000 | $197^{*}$ | $237^{*}$ | $236^{*}$ | $249^{*}$ | $236^{*}$ | $222^{*}$ |
| 2001 | $176^{*}$ | $219^{*}$ | $224^{*}$ | $257^{*}$ | $245^{*}$ | $237^{*}$ |
| 2002 | 295 | $313^{*}$ | $339^{*}$ | $367^{*}$ | $365^{*}$ | $371^{*}$ |
| 2003 | 156 | 183 | $135^{*}$ | $161^{*}$ | $171^{*}$ | $185^{*}$ |
| 2004 | 462 | 755 | 672 | $665^{*}$ | $668^{*}$ | $610^{*}$ |
| 2005 |  | 521 | 731 | 943 | $975^{*}$ | $1028^{*}$ |
| 2006 |  |  | 463 | 832 | 1036 | $811^{*}$ |
| 2007 |  |  |  | 202 | 208 | 212 |
| 2008 |  |  |  |  | 149 | 101 |
| 2009 |  |  |  |  |  | 303 |

### 4.4.4 Prediction data (Table 4.11, Table 4.19)

Weights at age and proportions mature at age show strong cyclic patterns related to periods of good recruitment. The Working Group believes that the estimated recruitment in the most recent years is so high that it will affect growth and maturation processes. The Working Group therefore decided to use similar trends in weight at age, maturity and natural mortality as has been observed in previous periods following good recruitment. The input data for making the prediction are presented in Table 4.19 :

- The estimated recruitment from RCT for 2010-2012 is given in Table 4.19.
- The assessment shows a decreasing trend in F from 2007 to 2009 and the F in 2009 is thus considered to be a better estimate for $F$ in the present year (2010) than using a three year average F.
- The average fishing pattern observed in the 3 last years, scaled to F status quo was used for distribution of fishing mortality at age for 2010-2012.
- Smoothed observed maturity for 2010 , smoothed average maturity for the 1982-1985, 1990-1993 and 2000-2006 year classes for 2011-2012.
- Smoothed observed weights at age in the stock for 2010, smoothed average weights for the 1982-1985, 1990-1993 and 2000-2006 year classes for 20112012.
- The average weights in the catch for the 1982-1985, 1990-1993 and 20002006 year classes for 2010-2012.
- Natural mortality - average for the 3 last years (2007-2009).
- Stock numbers and fishing mortalities from the standard VPA.


### 4.5 Results of the Assessments

### 4.5.1 Comparison of assessments

In case the IUU catches estimates equal zero this year, there are no comparisons the differences between assessment with and without IUU estimates as in previous reports.
The current assessment estimated the total stock to be about 5 \% lower and SSB 19 \% lower in 2008 compared to the previous assessment. F in 2008 is $4 \%$ higher than estimated last year.
There is a notable systematic difference between the time series of abundance at age from the XSA and those observed by the surveys, namely that the XSA time series is smoother and generally does not follow the relatively sharp peaks seen in the surveys. Neither the reason for this nor its significance for the assessment is fully understood (See the 2009 report, Figure 4.7). This investigation can be addressed at the benchmark.

### 4.5.2 Fishing mortality and VPA (Tables 4.12-4.18 and Figures 4.1A-D, 4.54.6)

The tuning diagnostics of the final XSA (predation included) is given in Table 4.12, the retrospective plot in Figure 4.5 and the log catchability residuals plot is presented in Figure 4.6.

The proportion of M and F before spawning was set to 0 . Fishing mortality are given in Table 4.13, while the stock numbers and spawning stock numbers, stock biomass at age and the spawning biomass at age of the final VPA are given in Tables 4.14-4.17. A summary of landings, fishing mortality, spawning stock biomass, and recruitment since 1950 are given in Table 4.18 and Figures 4.1A, 4.1B, 4.1C and 4.1D.

The assessments show a stable fishing mortality over the last three years, but the Fishing mortalities for the most recent years have been estimated higher this year than last year. Fishing mortality is currently decreasing and estimated well below the long term mean but only slightly below Fpa.

The dominating feature of the updated assessments is the rapid increase in biomass in 2009 and further in 2010, which is mainly the effect of a vastly improved recruitment. The increase in spawning stock biomass is still present but the rate of increase appears slightly bigger compared to last year.

### 4.5.3 Catch options for 2011-2012 (Tables 4.19-4.22)

Input to the predictions is given in Table 4.19. The estimated catch in 2009 gives $\mathrm{F}=0.31$ and the corresponding spawning stock biomasses is 285470 t at the beginning of 2010 , which is among the highest recorded.

The average F for the last three years ( F status quo) was used last year, but for 2010 it was decided use $F$ status quo equal $\mathrm{F}_{2009}=0.31$. Fishing pattern were calculated the same way as last year and taken as three year average scaled to $F$ status quo.
The deterministic projection shows a further increase in SSB in the beginning of 2011 (Table 4.20).
Fishing at $\mathrm{F}_{\mathrm{pa}}$ in 2011 corresponds to total landings about 330000 t , raising the SSB at the beginning of 2012 further to more than 450000 t (Table 4.21). But the $25 \%$ limita-
tion restricting the TAC (see Section 4.7.2) results in a TAC on 303000 t for 2011 ( $+25 \%$ compared to TAC for 2010 equal to 243000 t) predicting $F=0.31$ in 2011 (Table 4.22).

### 4.6 Comments to the assessment and forecasts

The problems using XSA on the Northeast Arctic haddock stock was discussed in 2008 (WD 24, AFWG 2008). The main conclusion was, and still is, that the XSA output is rather sensitive to the XSA settings (Figure 4.7), but the reasons for this are not fully understood.

The table below mainly reflects uncertainties in assessment and forecasts.

| SOURCE OF UNCERTAINTY | DESCRIPTION | Comments |
| :---: | :---: | :---: |
| Incomplete survey coverage (1) | Since 1997 all of the surveys used for tuning have been affected by an incomplete coverage for some of the years. (Due to Norwegian vessels not been given access to REZ, Russian vessels not been given access to NEZ). | All indices affected have been corrected using a factor based on geographical distributions observed before and after the incomplete coverage. This procedure is likely to introduce increased uncertainty to the indices (see AFWG 2007 and 4.2). |
| Incomplete survey coverage (2) | None of the surveys have a complete coverage of the stock. The proportion of a year class being outside the coverage varies between year classes (see also the WG report from 2002). | May appear as year class dependent changes in survey catchability. Catches of haddock in Norwegian statistical areas 06 and 07 (coastal areas) are added to the NEA haddock. These include haddock of older ages compared to the landings of NEA haddock. Since the surveys do not cover the coastal regions the coverage of older ages may be poorer. |
| Correlated error structures | Year effects in a survey are quite common. The year effect introduces correlated errors between the age groups, but in this case also between survey series. |  |
| Discards | The level of discarding is not known. | Discarding is known to be a (varying) problem in the longline fisheries related to the abundance of haddock close to, but below the minimum landing size. |
| Unreported catches | This year, estimates for unreported catches were provided for 2002-2008, but for 2009 set as zero. | The estimates were considered quite uncertain, but the uncertainty has decreased in recent years. |
| Predation on <br> young age <br> groups  | The survival due to predation (to a large extent by cod) varies substantially fromyear to year. | The predictions of young age groups are very uncertain, escpecially for the 3 -years HCR. |
| Sampling error | Estimation of catch at age is based on sampling of catches. The error in the estimates caused by sampling can be considerable even if the total catch is known. The estimation of the abundance indices from surveys will also be affected by sampling error. | The effect of not taking sampling error into account when fitting models to data may introduce bias in the resulting estimates. This bias is likely to increase with sampling error. |

### 4.7 Reference points and harvest control rules (Tables 4.23 and Figures 4.2-4.3)

### 4.7.1 Biomass and fishing mortality reference points

In 2006 the data used in the assessment were revised for the entire time series, and some additional catches previously not included into statistic (Norwegian statistical regions 06 and 07) have been added (see AFWG 2006 report (ICES 2006a) and WKHAD report (ICES 2006b) for a detailed description). The reference points have not been updated accordingly. The biomass reference points previously adopted and currently used by ACFM for this stock are $\boldsymbol{B}_{\lim }=50,000 \mathrm{t}$ and $\boldsymbol{B}_{\mathrm{pa}}=80,000 \mathrm{t}$. The fishing mortality reference points are $\mathbf{F}_{\mathrm{lim}}=0.49$ and $\mathbf{F}_{\mathrm{pa}}=0.35$ (Figure 4.4). Due to time constraints there was no work done during the AFWG meeting on revising the reference points of NEA haddock. The WG leave this w ork to the next benchmark assessment. A plot of SSB versus recruitment is shown in Figure 4.2. Yield and SSB per recruit (YPR and SPR) are presented in Table 4.23 and Figure4.3.

### 4.7.2 Harvest control rule

The harvest control rule (HCR) was evaluated by ICES in 2007 (ICES 2007a) and found to be in agreement with the precautionary approach. The agreed HCR for haddock is as follows (Protocol of the $36^{\text {th }}$ Session of The Joint Norwegian Russian Fishery Commission, 10 October 2007):

- TAC for the next year will be set at level corresponding to $F_{p a}$.
- The TAC should not be changed by more than +/- $25 \%$ compared with the previous year TAC.
- If the spawning stock falls below $B_{p a}$, the procedure for establishing TAC should be based on a fishing mortality that is linearly reduced from $F_{p a}$ at $B_{p a}$ to $F=0$ at $S S B$ equal to zero. At SSB-levels below $B_{p a}$ in any of the operational years (current year and a year ahead) there should be no limitations on the year-to-year variations in TAC.


### 4.8 Comments to Technical Minutes from ACFM

Our comments to Technical Minutes from ACFM are in italics below each comment from ACFM that requires a response.
General comments
The report is well done and the text is an update from last years report with relative changes. The assessment gives the increasing stock as a result of a reduction in fishing mortality and good recruitment last years. The Quality Handbook was revised.
Technical comments
The review was restricted to a check whether the procedures described in the technical annex (handbook) were applied. This was the case. No deviations were spotted.

The procedures used were the same as last year.
The values of stock weights have been changed in 1950-1984 and in 1985-2008 (see Chapter 4.3.3), estimates of consumption of NEA haddock by NEA cod (see chapter 4.4.2) and maturity at age (see chapter 4.3.5) were updated also.

The results of the assessment show that in case of haddock the XSA is rather sensitive to the XSA settings and shows large deviations from last year. The WG discussed it and gives uncertainties in assessment and forecasts in the report. The main uncertain-
ties derive from the biased catch statistics. There are no estimates of discarding. Both Russian (2006) and Norwegian (2007) bottom trawl surveys coverage were reduced compared to previous years.
The assessment indicates that the increasing of the SSB is relative with decreasing F and due to the high level of recruitment.
Decreasing of estimated IUU catches are explained in the Quality handbook. It should be placed in report.
This is now mentioned in the report.
The precautionary reference points are set based on an assessment carried out in 2000. The present assessment indicates that the historical biomasses estimates have been revised and that the technical basis for the biomass reference points is no longer valid. ICES needs to reconsider the PA reference points in a benchmark assessment in 2010.

A benchmark assessment is planned for 2011.
Remarks by the reviewer
In the report not mentioned why in 2006 was decided to include Norwegian landings of haddock from Norwegian statistical areas 06 ad 07 . Have to be referred where and when it was.

A reference to WKHAD 2006 is included.
Inspection of historical material raises questions for Norwegian statistical area 06 and 07 in table 4.1. The nominal catch for years 1960-1979 looks something erratically. Have to some explanations below table or in the text.
Due to uncertainties in the reliability of these catch statistics the WG decided to round off these numbers to the nearest thousand tonnes at the time of implementation in 2006. A footnote is added.
No details provided for the sampling data for length and age that are used for estimating catch at age. Is it enough or not?
The biological sampling of NEA haddock catches is considered to be fairly good. However, the present sampling is believed to be less precise because of the termination of a Norwegian sampling program in Q3 2009 (WD 6). A sentence is included in the text.
There are different estimates of unreported catches by Norway and Russian. This assumes to make 3 different assessments and prognoses, based on different assumptions on unreported landings and without it. This year was only 2 assessments with the highest and zero unreported catches estimates were made. Seems to be some explanations why.
This issue was discussed at the working group and it was agreed on using the Norwegian estimates for IUU. This conclusion is strengthened by the decision by ACFM to use the assessment including IUU as basis for the adtoice.
Retrospective runs for the 2000-2002 and middle 90 -th looks strange (figure 4.8). Such a retro needs additional investigation on next benchmark.

This is noted and will be investigated on the benchmark.
Residuals for the ages 7-8 both for all surveys are too high (figure 4.9) and not discussed in the report. Seems that data is not fully correct or incomplete.

This is noted and will be part of the next benchmark.
Have no clear explanation why the CPUE data don't used in the assessment.
Available CPUE data are not considered to be reliable enough to be used in assessment.
Conclusions
The assessment has been performed correctly. There is need for a benchmark in the short time. If a future benchmark the effect of different unreported (both IUU and discards) catches between years should be investigated. Surveys data to be revised again. Suggested to review data on weight at age matrix, seems that some problems with the age reading presented by different nations.

Comparative age readings between Norway and Russia have shown quite large discrepancies (up to 20\%) in the past. However, discrepancies have decreased in recent years and were below 10\% in 2009 (WD 14).

The present management plan is in accordance with a precautionary approach and the stock is harvested sustainable. However, unreported catches and discards are an important issue for this stock and reduce the effect of management measures and the objectives of the harvest control rule.
The problem of IUU is considered smaller in recent years and negligible in 2009.
The information given by the assessments is sufficient to provide advice.

### 4.9 Proposals to benchmark 2011.

1. Revising inputs and fleet data.
2. Revising reference points

3 Revise HCR
4 MSY approach addition

Table 4.1 Northeast Arctic haddock. Total nominal catch ( $t$ ) by fishing areas. (Data provided by Working Group members).

| Year | Sub- area I | Division Па | Division Ilb | Unre porte ${ }^{2}$ | Total ${ }^{3}$ | Norw. stat. <br> a reas 06 and $07^{4}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | 12502 | 27781 | 1844 | - | 154651 | 6000 |
| 1961 | 16515 | 25641 | 2427 | - | 193224 | 4000 |
| 1962 | 16056 | 25125 | 1723 | - | 187409 | 3000 |
| 1963 | 12433 | 20956 | 936 | - | 146224 | 4000 |
| 1964 | 79262 | 18784 | 1112 | - | 99158 | 6000 |
| 1965 | 98921 | 18719 | 943 | - | 118583 | 6000 |
| 1966 | 12500 | 35143 | 1626 | - | 161778 | 5000 |
| 1967 | 10799 | 27962 | 440 | - | 136398 | 3000 |
| 1968 | 14097 | 40031 | 725 | - | 181726 | 3000 |
| 1969 | 89948 | 40306 | 566 | - | 130820 | 2000 |
| 1970 | 60631 | 27120 | 507 | - | 88258 | - |
| 1971 | 56989 | 21453 | 463 | - | 78905 | - |
| 1972 | 22188 | 42111 | 2162 | - | 266153 |  |
| 1973 | 28564 | 23506 | 13077 | - | 322227 | - |
| 1974 | 15905 | 47037 | 15069 | - | 221157 | 10000 |
| 1975 | 12169 | 44337 | 9729 | - | 175758 | 6000 |
| 1976 | 94054 | 37562 | 5648 | - | 137264 | 2000 |
| 1977 | 72159 | 28452 | 9547 | - | 110158 | 2000 |
| 1978 | 63965 | 30478 | 979 | - | 95422 | 2000 |
| 1979 | 63841 | 39167 | 615 | - | 103623 | 6000 |
| 1980 | 54205 | 33616 | 68 | - | 87889 | 5098 |
| 1981 | 36834 | 39864 | 455 | - | 77153 | 4767 |
| 1982 | 17948 | 29005 | 2 | - | 46955 | 3335 |
| 1983 | 5837 | 16859 | 1904 | - | 24600 | 3112 |
| 1984 | 2934 | 16683 | 1328 | - | 20945 | 3803 |
| 1985 | 27982 | 14340 | 2730 | - | 45052 | 3583 |
| 1986 | 61729 | 29771 | 9063 | - | 100563 | 4021 |
| 1987 | 97091 | 41084 | 16741 | - | 154916 | 3194 |
| 1988 | 45060 | 49564 | 631 | - | 95255 | 3756 |
| 1989 | 29723 | 28478 | 317 | - | 58518 | 4701 |
| 1990 | 13306 | 13275 | 601 | - | 27182 | 2912 |
| 1991 | 17985 | 17801 | 430 | - | 36216 | 3045 |
| 1992 | 30884 | 28064 | 974 | - | 59922 | 5634 |
| 1993 | 46918 | 32433 | 3028 | - | 82379 | 5559 |
| 1994 | 76748 | 50388 | 8050 | - | 135186 | 6311 |
| 1995 | 75860 | 53460 | 13128 | - | 142448 | 5444 |
| 1996 | 11274 | 61722 | 3657 | - | 178128 | 5126 |
| 1997 | 78128 | 73475 | 2756 | - | 154359 | 5987 |
| 1998 | 45640 | 53936 | 1054 | - | 100630 | 6338 |
| 1999 | 38291 | 40819 | 4085 | - | 83195 | 5743 |
| 2000 | 25931 | 39169 | 3844 | - | 68944 | 4536 |
| 2001 | 35072 | 47245 | 7323 | - | 89640 | 4542 |
| 2002 | 40721 | 42774 | 12567 | 18736/5310 | 114798/101372 | 6898 |
| 2003 | 53653 | 43564 | 8483 | 33226/9417 | 138926/115117 | 4279 |
| 2004 | 64873 | 47483 | 12146 | 33777/8661 | 158279/133163 | 3743 |
| 2005 | 53518 | 48081 | 16416 | 40283/9949 | 158298/127964 | 5538 |
| 2006 | 51124 | 47291 | 33291 | 21451/8949 | 153157/140655 | 5410 |
| 2007 | 62904 | 58141 | 25927 | 14553/3102 | 161525/150074 | 7110 |
| 2008 | 58379 | 60178 | 31219 | 5828/- | 155604/149776 | 6629 |
| $2009{ }^{1}$ | 58177 | 66065 | 76270 | 0 | 200512 | 4498 |

1 Provisional figures, Norwegian catches on Russian quotas are included
2 Figures based on Norwegian/Russian IUU estimates. From 2009, IUU estimates are made by a Joint Russian-Norwegian analysis group under the Russian-Norwegian Fisheries Commission.
3 Figures based on Norwegian/Russian IUU estimates. During the period 2002-2008, the Norwegian IUU-estimates were included in the final assessments
4 Included in total landings and in landings in region IIa, catches prior to 1980 is rounded off due to uncertainties in reliability.

Table 4.2 Northeast Arctic haddock. Total nominal catch (' 000 t) by trawl and other gear for each area.

|  | Sub-area I |  | Division IIa |  | Division Ilb |  | Unre porte ${ }^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Trawl | Others | Traw | Others | Traw | Others |  |
| 1967 | 73.7 | 34.3 | 20.5 | 7.5 | 0.4 | - |  |
| 1968 | 98.1 | 42.9 | 31.4 | 8.6 | 0.7 | - |  |
| 1969 | 41.4 | 47.8 | 33.2 | 7.1 | 1.3 | - |  |
| 1970 | 37.4 | 23.2 | 20.6 | 6.5 | 0.5 |  |  |
| 1971 | 27.5 | 29.2 | 15.1 | 6.7 | 0.4 | - |  |
| 1972 | 193.9 | 27.9 | 34.5 | 7.6 | 2.2 | - |  |
| 1973 | 242.9 | 42.8 | 14.0 | 9.5 | 13.1 | - |  |
| 1974 | 133.1 | 25.9 | 39.9 | 7.1 | 15.1 | - |  |
| 1975 | 103.5 | 18.2 | 34.6 | 9.7 | 9.7 | - |  |
| 1976 | 77.7 | 16.4 | 28.1 | 9.5 | 5.6 | - |  |
| 1977 | 57.6 | 14.6 | 19.9 | 8.6 | 9.5 | - |  |
| 1978 | 53.9 | 10.1 | 15.7 | 14.8 | 1.0 | - |  |
| 1979 | 47.8 | 16.0 | 20.3 | 18.9 | 0.6 | - |  |
| 1980 | 30.5 | 23.7 | 14.8 | 18.9 | 0.1 | - |  |
| 1981 | 18.8 | 17.7 | 21.6 | 18.5 | 0.5 | - |  |
| 1982 | 11.6 | 11.5 | 23.9 | 13.5 |  | - |  |
| 1983 | 3.6 | 2.2 | 8.7 | 8.2 | 0.2 | 1.7 |  |
| 1984 | 1.6 | 1.3 | 7.6 | 9.1 | 0.1 | 1.2 |  |
| 1985 | 24.4 | 3.5 | 6.2 | 8.1 | 0.1 | 2.6 |  |
| 1986 | 51.7 | 10.1 | 14.0 | 15.8 | 0.8 | 8.3 |  |
| 1987 | 79.0 | 18.1 | 23.0 | 18.1 | 3.0 | 13.8 |  |
| 1988 | 28.7 | 16.4 | 34.3 | 15.3 | 0.6 | 0.0 |  |
| 1989 | 20.0 | 9.7 | 13.5 | 15.0 | 0.3 | 0.0 |  |
| 1990 | 4.4 | 8.9 | 5.1 | 8.2 | 0.6 | 0.0 |  |
| 1991 | 9.0 | 8.9 | 8.9 | 8.9 | 0.2 | 0.2 |  |
| 1992 | 21.3 | 9.6 | 11.9 | 16.1 | 1.0 | 0.0 |  |
| 1993 | 35.3 | 11.6 | 14.5 | 17.9 | 3.0 | 0.0 |  |
| 1994 | 58.6 | 18.2 | 26.1 | 24.3 | 7.9 | 0.2 |  |
| 1995 | 63.9 | 12.0 | 29.6 | 23.8 | 12.1 | 1.0 |  |
| 1996 | 98.3 | 14.4 | 36.5 | 25.2 | 3.4 | 0.3 |  |
| 1997 | 57.4 | 20.7 | 44.9 | 28.6 | 2.5 | 0.3 |  |
| 1998 | 26.0 | 19.6 | 27.1 | 26.9 | 0.7 | 0.3 |  |
| 1999 | 29.4 | 8.9 | 19.1 | 21.8 | 4.0 | 0.1 |  |
| 2000 | 20.1 | 5.9 | 18.8 | 20.4 | 3.7 | 0.1 |  |
| 2001 | 28.4 | 6.7 | 23.4 | 23.8 | 7.0 | 0.3 |  |
| 2002 | 30.5 | 10.2 | 19.5 | 23.3 | 12.5 | 0.1 | 18.7/5.3 |
| 2003 | 42.7 | 10.9 | 21.9 | 21.7 | 8.1 | 0.4 | 33.2/9.4 |
| 2004 | 52.4 | 12.5 | 27.0 | 20.5 | 11.5 | 0.6 | 33.8/8.7 |
| 2005 | 38.5 | 15.0 | 24.9 | 20.9 | 13.0 | 1.6 | 40.3/9. |
| 2006 | 40.1 | 11 | 22 | 25.3 | 30.1 | 3.2 | 21.5/8. |
| 2007 | 51.8 | 11.1 | 30.5 | 27.7 | 20.4 | 5.5 | 14.6/3.1 |
| 2008 | 46.8 | 11.6 | 30.9 | 29.3 | 24.9 | 6.3 | 5.8/- |
| $2009{ }^{1}$ | 49.0 | 8.8 | 40.1 | 25.3 | 67.1 | 7.8 |  |

Table 4.3 Northeast Arctic haddock. Nominal catch (t) by countries. Sub-area I and Divisions IIa and IIb combined. (Data provided by Working Group members).

| Year | Faroe Islands | France | German Dem.Re. | Fed. Re. Germ. | Norway ${ }^{4}$ | Poland | Unite d Kingdom | Russia ${ }^{2}$ | Others | Unre porte d ca tches ${ }^{3}$ | Total ${ }^{3}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | 172 | - | - | 5597 | 46263 | - | 45469 | 57025 | 125 | - | 154651 |
| 1961 | 285 | 220 | - | 6304 | 60862 | - | 39650 | 85345 | 558 | - | 193224 |
| 1962 | 83 | 409 | - | 2895 | 54567 | - | 37486 | 91910 | 58 | - | 187408 |
| 1963 | 17 | 363 | - | 2554 | 59955 | - | 19809 | 63526 | - | - | 146224 |
| 1964 | - | 208 | - | 1482 | 38695 | - | 14653 | 43870 | 250 | - | 99158 |
| 1965 | - | 226 | - | 1568 | 60447 | - | 14345 | 41750 | 242 | - | 118578 |
| 1966 | - | 1072 | 11 | 2098 | 82090 | - | 27723 | 48710 | 74 | - | 161778 |
| 1967 | - | 1208 | 3 | 1705 | 51954 | - | 24158 | 57346 | 23 | - | 136397 |
| 1968 | - | - | - | 1867 | 64076 | - | 40129 | 75654 | - | - | 181726 |
| 1969 | 2 | - | 309 | 1490 | 67549 | - | 37234 | 24211 | 25 | - | 130820 |
| 1970 | 541 | - | 656 | 2119 | 37716 | - | 20423 | 26802 | - | - | 88257 |
| 1971 | 81 | - | 16 | 896 | 45715 | 43 | 16373 | 15778 | 3 | - | 78905 |
| 1972 | 137 | - | 829 | 1433 | 46700 | 1433 | 17166 | 196224 | 2231 | - | 266153 |
| 1973 | 1212 | 3214 | 22 | 9534 | 86767 | 34 | 32408 | 186534 | 2501 | - | 322226 |
| 1974 | 925 | 3601 | 454 | 23409 | 66164 | 3045 | 37663 | 78548 | 7348 | - | 221157 |
| 1975 | 299 | 5191 | 437 | 15930 | 55966 | 1080 | 28677 | 65015 | 3163 | - | 175758 |
| 1976 | 536 | 4459 | 348 | 16660 | 49492 | 986 | 16940 | 42485 | 5358 | - | 137264 |
| 1977 | 213 | 1510 | 144 | 4798 | 40118 | - | 10878 | 52210 | 287 | - | 110158 |
| 1978 | 466 | 1411 | 369 | 1521 | 39955 | 1 | 5766 | 45895 | 38 | - | 95422 |
| 1979 | 343 | 1198 | 10 | 1948 | 66849 | 2 | 6454 | 26365 | 454 | - | 103623 |
| 1980 | 497 | 226 | 15 | 1365 | 66501 | - | 2948 | 20706 | 246 | - | 92504 |
| 1981 | 381 | 414 | 22 | 2402 | 63435 | Spain | 1682 | 13400 | - | - | 81736 |
| 1982 | 496 | 53 | - | 1258 | 43702 | - | 827 | 2900 | - | - | 49236 |
| 1983 | 428 | - | 1 | 729 | 22364 | 139 | 259 | 680 | - | - | 24600 |
| 1984 | 297 | 15 | 4 | 400 | 18813 | 37 | 276 | 1103 | - | - | 20945 |
| 1985 | 424 | 21 | 20 | 395 | 21272 | 77 | 153 | 22690 | - | - | 45052 |
| 1986 | 893 | 12 | 75 | 1079 | 52313 | 22 | 431 | 45738 | - | - | 100563 |
| 1987 | 464 | 7 | 83 | 3105 | 72419 | 59 | 563 | 78211 | 5 | - | 154916 |


| 1988 | 1113 | 116 | 78 | 1323 | 60823 | 72 | 435 | 31293 | 2 | - | 95255 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 1217 | - | 26 | 171 | 36451 | 1 | 590 | 20062 | - | - | 58518 |
| 1990 | 705 | - | 5 | 167 | 20621 | - | 494 | 5190 | - | - | 27182 |
| 1991 | 1117 | - | Greenld | 213 | 22178 | - | 514 | 12177 | 17 | - | 36216 |
| 1992 | 1093 | 151 | 1719 | 387 | 36238 | 38 | 596 | 19699 | 1 | - | 59922 |
| 1993 | 546 | 1215 | 880 | 1165 | 40978 | 76 | 1802 | 35071 | 646 | - | 82379 |
| 1994 | 2761 | 678 | 770 | 2412 | 71171 | 22 | 4673 | 51822 | 877 | - | 135186 |
| 1995 | 2833 | 598 | 1097 | 2675 | 76886 | 14 | 3111 | 54516 | 718 | - | 142448 |
| 1996 | 3743 | 6 | 1510 | 942 | 94527 | 669 | 2275 | 74239 | 217 | - | 178128 |
| 1997 | 3327 | 540 | 1877 | 972 | 103407 | 364 | 2340 | 41228 | 304 | - | 154359 |
| 1998 | 1903 | 241 | 854 | 385 | 75108 | 257 | 1229 | 20559 | 94 | - | 100630 |
| 1999 | 1913 | 64 | 437 | 641 | 48182 | 652 | 694 | 30520 | 92 | - | 83195 |
| 2000 | 631 | 178 | 432 | 880 | 42009 | 502 | 747 | 22738 | 827 | - | 68944 |
| 2001 | 1210 | 324 | 553 | 554 | 49067 | 1497 | 1068 | 34307 | 1060 | - | 89640 |
| 2002 | 1564 | 297 | 858 | 627 | 52247 | 1505 | 1125 | 37157 | 682 | 18736/5310 | 114798/101372 |
| 2003 | 1959 | 382 | 1363 | 918 | 56485 | 1330 | 1018 | 41142 | 1103 | 33226/9417 | 138926/115117 |
| 2004 | 2484 | 103 | 1680 | 823 | 62192 | 54 | 1250 | 54347 | 1569 | 33777/8661 | 158279/133163 |
| 2005 | 2138 | 333 | 15 | 996 | 60850 | 963 | 1899 | 50012 | 1262 | 40283/9949 | 158751/128417 |
| 2006 | 2390 | 883 | 1830 | 989 | 69272 | 703 | 1164 | 53313 | 1162 | 21451/8949 | 153157/140/655 |
| 2007 | 2307 | 277 | 1464 | 1123 | 71244 | 125 | 1351 | 66569 | 2511 | 14553/3102 | 161525/150074 |
| 2008 | 2687 | 311 | 1659 | 535 | 72779 | 283 | 971 | 68792 | 1759 | 5828/- | 155604/149776 |
| $2009{ }^{1}$ | 2953 | 529 | 1407 | 1942 | 104354 | 317 | 1315 | 85514 | 2181 | 0 | 200512 |

1 Provisional figures.
3 Figures based on Norwegian/Russian IUU estimates
4 included landings in Norwegian statistical areas 06 and 07 (from 1983)

Table 4.4. Northeast Arctic haddock. Catch numbers at age (numbers, '000)

| age | 1950 | 1951 | 1952 | 1953 | 1954 | 1955 | 1956 | 1957 | 1958 | 1959 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 3189 | 65643 | 6012 | 64528 | 6563 | 1154 | 16437 | 2074 | 1727 | 20318 |
| 4 | 37949 | 9178 | 151996 | 13013 | 154696 | 10689 | 5922 | 24704 | 5914 | 7826 |
| 5 | 35344 | 18014 | 13634 | 70781 | 5885 | 176678 | 14713 | 7942 | 31438 | 7243 |
| 6 | 18849 | 13551 | 9850 | 5431 | 27590 | 4993 | 127879 | 12535 | 5820 | 14040 |
| 7 | 28868 | 6808 | 4693 | 2867 | 3233 | 28273 | 3182 | 46619 | 12748 | 3154 |
| 8 | 9199 | 6850 | 3237 | 1080 | 1302 | 1445 | 8003 | 1087 | 17565 | 2237 |
| 9 | 1979 | 3322 | 2434 | 424 | 712 | 271 | 450 | 1971 | 822 | 5918 |
| 10 | 1093 | 1182 | 606 | 315 | 319 | 100 | 200 | 356 | 1072 | 285 |
| 11+ | 2977 | 1348 | 880 | 1005 | 543 | 100 | 185 | 176 | 601 | 500 |
| TOTNU | 139447 | 125896 | 193342 | 159444 | 200843 | 223703 | 176971 | 97464 | 77707 | 61521 |
| TONS | 132125 | 120077 | 127660 | 123920 | 156788 | 202286 | 213924 | 123583 | 112672 | 88211 |
| SOPCOF\% year | 61 | 80 | 56 | 68 | 66 | 64 | 77 | 78 | 87 | 104 |
| age | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 |
| 3 | 39910 | 15429 | 39503 | 28466 | 22363 | 5936 | 26345 | 15907 | 657 | 1524 |
| 4 | 70912 | 56855 | 30868 | 72736 | 49290 | 46356 | 22631 | 41346 | 67632 | 1968 |
| 5 | 13647 | 63351 | 48903 | 18969 | 30672 | 40201 | 63176 | 13496 | 41267 | 44634 |
| 6 | 7101 | 8706 | 33836 | 13579 | 5815 | 12631 | 29048 | 25719 | 7748 | 19002 |
| 7 | 6236 | 3578 | 3201 | 9257 | 3527 | 1679 | 5752 | 8872 | 15599 | 3620 |
| 8 | 1579 | 4407 | 1341 | 1239 | 2716 | 974 | 582 | 1616 | 5292 | 4937 |
| 9 | 2340 | 788 | 1773 | 559 | 833 | 897 | 438 | 218 | 655 | 1628 |
| 10 | 2005 | 527 | 242 | 409 | 104 | 123 | 189 | 175 | 182 | 316 |
| 11+ | 606 | 1434 | 756 | 375 | 633 | 802 | 242 | 271 | 286 | 109 |
| TOTNU | 144336 | 155075 | 160423 | 145589 | 115953 | 109599 | 148403 | 107620 | 139318 | 77738 |
| TONS | 154651 | 193224 | 187408 | 146224 | 99158 | 118578 | 161778 | 136397 | 181726 | 130820 |
| SOPCOF\% year | 94 | 98 | 93 | 85 | 72 | 85 | 84 | 98 | 98 | 111 |
| age | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| 3 | 23444 | 1978 | 230942 | 70679 | 9685 | 10037 | 13994 | 55967 | 47311 | 17540 |
| 4 | 2454 | 24358 | 22315 | 260520 | 41706 | 14088 | 13454 | 22043 | 18812 | 35290 |
| 5 | 1906 | 1257 | 42981 | 24180 | 88120 | 33871 | 6810 | 7368 | 4076 | 10645 |
| 6 | 22417 | 918 | 3206 | 6919 | 5829 | 49711 | 20796 | 2586 | 1389 | 1429 |
| 7 | 8100 | 9279 | 1611 | 422 | 4138 | 2135 | 40057 | 7781 | 1626 | 812 |
| 8 | 2012 | 3056 | 6758 | 426 | 382 | 1236 | 1247 | 11043 | 2596 | 546 |
| 9 | 2016 | 826 | 2638 | 1692 | 618 | 92 | 1350 | 311 | 6215 | 1466 |
| 10 | 740 | 1043 | 900 | 529 | 2043 | 131 | 193 | 388 | 162 | 2310 |
| 11+ | 293 | 534 | 1652 | 584 | 1870 | 934 | 1604 | 379 | 400 | 323 |
| TOTNU | 63382 | 43249 | 313003 | 365951 | 154391 | 112235 | 99505 | 107866 | 82587 | 70361 |
| TONS | 88257 | 78905 | 266153 | 322226 | 221157 | 175758 | 137264 | 110158 | 95422 | 103623 |
| SOPCOF\% | 100 | 128 | 90 | 84 | 109 | 109 | 87 | 90 | 106 | 127 |

Table 4.4 (continued).

| age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 627 | 486 | 883 | 1173 | 1271 | 29624 | 23113 | 5031 | 1439 | 2157 |
| 4 | 22878 | 2561 | 900 | 2636 | 1019 | 1695 | 68429 | 87170 | 12478 | 4986 |
| 5 | 21794 | 22124 | 3372 | 1360 | 1899 | 564 | 1565 | 64556 | 47890 | 16071 |
| 6 | 2971 | 10685 | 12203 | 2394 | 657 | 1009 | 783 | 960 | 20429 | 25313 |
| 7 | 250 | 1034 | 2625 | 2506 | 950 | 943 | 896 | 597 | 397 | 3198 |
| 8 | 504 | 162 | 344 | 1799 | 2619 | 886 | 393 | 376 | 178 | 147 |
| 9 | 230 | 162 | 75 | 267 | 352 | 1763 | 702 | 212 | 74 | 1 |
| 10 | 842 | 72 | 80 | 37 | 87 | 588 | 1144 | 230 | 88 | 28 |
| 11+ | 1460 | 963 | 649 | 292 | 77 | 281 | 987 | 738 | 446 | 177 |
| TOTNU | 51556 | 38249 | 21131 | 12464 | 8931 | 37353 | 98012 | 159870 | 83419 | 52078 |
| TONS | 87889 | 77153 | 46955 | 24600 | 20945 | 45052 | 100563 | 154916 | 95255 | 58518 |
| SOPCOF\% | 129 | 136 | 135 | 95 | 95 | 102 | 95 | 101 | 100 | 102 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 3 | 1015 | 4421 | 11571 | 13487 | 3374 | 2003 | 1662 | 2280 | 1701 | 16839 |
| 4 | 2580 | 3564 | 11567 | 19457 | 47821 | 16109 | 6818 | 5633 | 11304 | 8039 |
| 5 | 2142 | 2416 | 4099 | 13704 | 36333 | 72644 | 36473 | 12603 | 9258 | 15365 |
| 6 | 4046 | 3299 | 2642 | 4103 | 13264 | 19145 | 73579 | 32832 | 8633 | 6073 |
| 7 | 6221 | 4633 | 2894 | 1747 | 2057 | 6417 | 13426 | 49478 | 13801 | 4466 |
| 8 | 840 | 3953 | 3327 | 1886 | 903 | 746 | 2944 | 5636 | 19469 | 6355 |
| 9 | 134 | 461 | 3498 | 2105 | 1453 | 361 | 573 | 778 | 2113 | 6204 |
| 10 | 42 | 83 | 486 | 1965 | 2769 | 770 | 365 | 245 | 330 | 647 |
| 11+ | 71 | 54 | 84 | 323 | 2110 | 1576 | 1897 | 748 | 490 | 446 |
| TOTNU | 17091 | 22884 | 40168 | 58777 | 110084 | 119771 | 137737 | 110233 | 67099 | 64434 |
| TONS | 27182 | 36216 | 59922 | 82379 | 135186 | 142448 | 178128 | 154359 | 100630 | 83195 |
| SOPCOF\% | 98 | 96 | 102 | 100 | 99 | 98 | 98 | 95 | 99 | 98 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 3 | 1520 | 12971 | 7132 | 6803 | 7993 | 11452 | 4539 | 30707 | 14536 | 15379 |
| 4 | 29986 | 5230 | 46335 | 31448 | 21116 | 19369 | 35040 | 15213 | 44192 | 55013 |
| 5 | 6496 | 32049 | 11084 | 56480 | 41310 | 22887 | 27571 | 45992 | 15926 | 52498 |
| 6 | 5149 | 5279 | 21985 | 11736 | 41226 | 37067 | 15033 | 18516 | 31173 | 13679 |
| 7 | 2406 | 2941 | 2602 | 14541 | 4939 | 24461 | 16023 | 10642 | 9145 | 15382 |
| 8 | 1657 | 1137 | 1602 | 1637 | 4914 | 2393 | 8567 | 7889 | 4520 | 3800 |
| 9 | 1570 | 1161 | 482 | 2178 | 598 | 2997 | 1259 | 2570 | 2846 | 1669 |
| 10 | 1744 | 1169 | 448 | 858 | 1252 | 990 | 1298 | 678 | 1181 | 887 |
| 11+ | 437 | 1204 | 1029 | 1219 | 901 | 1524 | 718 | 988 | 654 | 960 |
| TOTNU | 50965 | 63141 | 92699 | 126900 | 124249 | 123140 | 110048 | 133195 | 124173 | 159267 |
| TONS | 68944 | 89640 | 114798 | 138926 | 158279 | 158298 | 153157 | 161525 | 155604 | 200512 |
| SOPCOF\% | 97 | 101 | 99 | 98 | 98 | 100 | 101 | 101 | 101 | 100 |

## Table 4.5. Northeast Arctic haddock. Catch weights at age (kg)

|  | 1950 | 1951 | 1952 | 1953 | 1954 | 1955 | 1956 | 1957 | 1958 | 1959 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 |
| 4 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 |
| 5 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 |
| 6 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 |
| 7 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 |
| 8 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 |
| 9 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 |
| 10 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 |
| 11+ | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 |
| $\begin{gathered} \text { SOPCOF\% } \\ \text { year } \end{gathered}$ | 61 | 80 | 56 | 68 | 66 | 64 | 77 | 78 | 87 | 104 |
|  |  |  |  |  |  |  |  |  |  |  |
|  | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 |
| 3 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 |
| 4 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 |
| 5 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 |
| 6 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 |
| 7 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 |
| 8 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 |
| 9 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 |
| 10 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 |
| 11+ | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 |
| SOPCOF\% year | 94 | 98 | 93 | 85 | 72 | 85 | 84 | 98 | 98 | 111 |
|  | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| 3 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 | 0.768 |
| 4 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 | 1.065 |
| 5 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 | 1.353 |
| 6 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 | 1.663 |
| 7 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 | 1.921 |
| 8 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 | 2.183 |
| 9 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 | 2.463 |
| 10 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 | 2.752 |
| 11+ | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 | 3.177 |
| SOPCOF\% | 100 | 128 | 90 | 84 | 109 | 109 | 87 | 90 | 106 | 127 |

Table 4.5 (continued).

|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.768 | 0.768 | 0.768 | 1.033 | 1.218 | 0.835 | 0.612 | 0.497 | 0.55 | 0.684 |
| 4 | 1.065 | 1.065 | 1.065 | 1.408 | 1.632 | 1.29 | 1.064 | 0.765 | 0.908 | 0.84 |
| 5 | 1.353 | 1.353 | 1.353 | 1.71 | 2.038 | 1.816 | 1.539 | 1.179 | 1.097 | 0.998 |
| 6 | 1.663 | 1.663 | 1.663 | 2.149 | 2.852 | 2.174 | 1.944 | 1.724 | 1.357 | 1.176 |
| 7 | 1.921 | 1.921 | 1.921 | 2.469 | 2.845 | 2.301 | 2.362 | 2.135 | 1.537 | 1.546 |
| 8 | 2.183 | 2.183 | 2.183 | 2.748 | 3.218 | 2.835 | 2.794 | 2.551 | 1.704 | 1.713 |
| 9 | 2.463 | 2.463 | 2.463 | 3.069 | 3.605 | 3.253 | 3.25 | 3.009 | 2.403 | 1.949 |
| 10 | 2.752 | 2.752 | 2.752 | 3.687 | 4.065 | 3.721 | 3.643 | 3.414 | 2.403 | 2.14 |
| 11+ | 3.177 | 3.177 | 3.177 | 4.516 | 4.667 | 4.416 | 5.283 | 4.213 | 2.571 | 2.685 |
| SOPCOF\% <br> year | 129 | 136 | 135 | 95 | 95 | 102 | 95 | 101 | 100 | 102 |
|  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 3 | 0.793 | 0.941 | 0.906 | 0.94 | 0.614 | 0.739 | 0.683 | 0.682 | 0.748 | 0.826 |
| 4 | 1.172 | 1.281 | 1.263 | 1.204 | 0.906 | 0.808 | 0.868 | 1.028 | 0.974 | 1.079 |
| 5 | 1.397 | 1.556 | 1.535 | 1.487 | 1.287 | 1.107 | 1.045 | 1.151 | 1.262 | 1.261 |
| 6 | 1.624 | 1.797 | 1.747 | 1.748 | 1.602 | 1.556 | 1.363 | 1.369 | 1.433 | 1.485 |
| 7 | 1.885 | 2.044 | 2.043 | 1.994 | 1.968 | 1.838 | 1.71 | 1.637 | 1.641 | 1.634 |
| 8 | 2.112 | 2.079 | 2.2 | 2.237 | 2.059 | 2.234 | 1.886 | 1.856 | 1.863 | 1.798 |
| 9 | 2.653 | 2.311 | 2.298 | 2.417 | 2.39 | 2.416 | 2.214 | 2.073 | 2.069 | 2.032 |
| 10 | 3.102 | 2.788 | 2.494 | 2.654 | 2.545 | 2.602 | 2.37 | 2.5 | 2.335 | 2.237 |
| 11+ | 3.338 | 3.219 | 2.652 | 3.026 | 2.893 | 3.13 | 2.675 | 2.554 | 2.81 | 2.712 |
| SOPCOF\% <br> year | 98 | 96 | 102 | 100 | 99 | 98 | 98 | 95 | 99 | 98 |
|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 3 | 0.853 | 0.751 | 0.687 | 0.594 | 0.636 | 0.722 | 0.745 | 0.652 | 0.658 | 0.706 |
| 4 | 1.186 | 1.104 | 1.001 | 0.875 | 0.886 | 0.906 | 1.041 | 0.899 | 0.901 | 1.023 |
| 5 | 1.395 | 1.459 | 1.363 | 1.113 | 1.183 | 1.121 | 1.287 | 1.197 | 1.242 | 1.279 |
| 6 | 1.588 | 1.709 | 1.643 | 1.364 | 1.508 | 1.343 | 1.504 | 1.435 | 1.515 | 1.536 |
| 7 | 1.808 | 1.921 | 1.975 | 1.361 | 1.821 | 1.619 | 1.72 | 1.722 | 1.781 | 1.807 |
| 8 | 1.989 | 2.182 | 2.086 | 1.972 | 2.075 | 2.036 | 2.082 | 1.99 | 2.18 | 2.11 |
| 9 | 2.264 | 2.331 | 2.294 | 1.636 | 2.339 | 2.177 | 2.377 | 2.309 | 2.33 | 2.406 |
| 10 | 2.415 | 2.609 | 2.487 | 1.877 | 2.58 | 2.382 | 2.738 | 2.715 | 2.664 | 2.532 |
| 11+ | 2.892 | 2.981 | 2.778 | 2.409 | 2.991 | 2.768 | 3.212 | 3.028 | 3.328 | 3.172 |
| SOPCOF\% | 97 | 101 | 99 | 98 | 98 | 100 | 101 | 101 | 101 | 100 |

## Table 4.6. Northeast Arctic haddock. Stock weights at age (kg)

|  | 1950 | 1951 | 1952 | 1953 | 1954 | 1955 | 1956 | 1957 | 1958 | 1959 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 |
| 4 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 |
| 5 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 |
| 6 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 |
| 7 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 |
| 8 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 |
| 9 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 |
| 10 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 |
| 11+ | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 |
| year |  |  |  |  |  |  |  |  |  |  |
|  | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 |
| 3 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 |
| 4 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 |
| 5 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 |
| 6 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 |
| 7 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 |
| 8 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 |
| 9 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 |
| 10 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 |
| 11+ | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 |
| year |  |  |  |  |  |  |  |  |  |  |
|  | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| 3 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 |
| 4 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 | 0.648 |
| 5 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 | 1.015 |
| 6 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 | 1.431 |
| 7 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 | 1.883 |
| 8 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 | 2.357 |
| 9 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 | 2.832 |
| 10 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 | 3.299 |
| 11+ | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 | 3.751 |

Table 4.6 (continued).

|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.457 | 0.627 | 0.633 | 0.546 | 0.369 | 0.358 | 0.302 | 0.315 | 0.351 | 0.41 |
| 4 | 0.844 | 0.819 | 1.101 | 1.12 | 0.973 | 0.672 | 0.656 | 0.56 | 0.584 | 0.644 |
| 5 | 1.084 | 1.285 | 1.251 | 1.651 | 1.692 | 1.48 | 1.04 | 1.023 | 0.885 | 0.921 |
| 6 | 1.555 | 1.509 | 1.769 | 1.725 | 2.239 | 2.311 | 2.034 | 1.454 | 1.441 | 1.261 |
| 7 | 2.096 | 2.009 | 1.958 | 2.271 | 2.219 | 2.833 | 2.945 | 2.607 | 1.894 | 1.89 |
| 8 | 2.76 | 2.557 | 2.469 | 2.417 | 2.773 | 2.713 | 3.412 | 3.572 | 3.176 | 2.345 |
| 9 | 3.095 | 3.195 | 3.01 | 2.925 | 2.872 | 3.261 | 3.193 | 3.96 | 4.173 | 3.726 |
| 10 | 3.462 | 3.525 | 3.612 | 3.448 | 3.367 | 3.316 | 3.725 | 3.65 | 4.468 | 4.739 |
| 11+ | 3.84 | 3.877 | 3.933 | 4.008 | 3.864 | 3.79 | 3.741 | 4.161 | 4.077 | 4.931 |
| year |  |  |  |  |  |  |  |  |  |  |
|  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 3 | 0.382 | 0.385 | 0.335 | 0.264 | 0.247 | 0.276 | 0.289 | 0.309 | 0.33 | 0.328 |
| 4 | 0.741 | 0.693 | 0.7 | 0.614 | 0.492 | 0.461 | 0.515 | 0.538 | 0.571 | 0.609 |
| 5 | 1.004 | 1.141 | 1.072 | 1.085 | 0.959 | 0.779 | 0.734 | 0.819 | 0.851 | 0.897 |
| 6 | 1.31 | 1.413 | 1.588 | 1.496 | 1.518 | 1.351 | 1.112 | 1.053 | 1.172 | 1.213 |
| 7 | 1.673 | 1.736 | 1.854 | 2.06 | 1.947 | 1.978 | 1.772 | 1.478 | 1.405 | 1.563 |
| 8 | 2.355 | 2.105 | 2.182 | 2.312 | 2.54 | 2.407 | 2.449 | 2.206 | 1.863 | 1.78 |
| 9 | 2.792 | 2.823 | 2.547 | 2.638 | 2.772 | 3.016 | 2.865 | 2.918 | 2.641 | 2.257 |
| 10 | 4.246 | 3.228 | 3.283 | 2.989 | 3.092 | 3.226 | 3.477 | 3.31 | 3.375 | 3.068 |
| 11+ | 5.263 | 4.73 | 3.644 | 3.728 | 3.422 | 3.538 | 3.667 | 3.916 | 3.736 | 3.811 |
| year |  |  |  |  |  |  |  |  |  |  |
|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 3 | 0.282 | 0.286 | 0.268 | 0.272 | 0.292 | 0.31 | 0.322 | 0.329 | 0.291 | 0.291 |
| 4 | 0.604 | 0.525 | 0.53 | 0.499 | 0.507 | 0.542 | 0.573 | 0.593 | 0.606 | 0.54 |
| 5 | 0.955 | 0.946 | 0.83 | 0.835 | 0.79 | 0.803 | 0.856 | 0.902 | 0.931 | 0.951 |
| 6 | 1.273 | 1.352 | 1.336 | 1.183 | 1.188 | 1.128 | 1.146 | 1.219 | 1.279 | 1.317 |
| 7 | 1.609 | 1.681 | 1.782 | 1.757 | 1.57 | 1.572 | 1.498 | 1.523 | 1.615 | 1.689 |
| 8 | 1.976 | 2.026 | 2.108 | 2.231 | 2.193 | 1.976 | 1.975 | 1.889 | 1.921 | 2.031 |
| 9 | 2.166 | 2.401 | 2.453 | 2.542 | 2.685 | 2.633 | 2.392 | 2.385 | 2.289 | 2.327 |
| 10 | 2.652 | 2.556 | 2.828 | 2.879 | 2.973 | 3.135 | 3.067 | 2.806 | 2.793 | 2.689 |
| 11+ | 3.479 | 3.04 | 2.942 | 3.25 | 3.297 | 3.395 | 3.573 | 3.486 | 3.213 | 3.191 |

Table 4.7. Northeast Arctic haddock. Natural mortality (M) at age

|  | 1950-1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.328 | 0.2074 | 0.2 | 0.6481 | 0.2 | 0.4052 | 0.2 | 0.3197 | 0.2 | 0.2059 |
| 4 | 0.2322 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 5 | 0.2238 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2027 | 0.2 | 0.2 | 0.2 | 0.2 |
| 6 | 0.2073 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 7 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 8 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 9 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 10 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 11+ | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| year |  |  |  |  |  |  |  |  |  |  |
|  | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| 3 | 0.2619 | 0.2962 | 0.3451 | 0.7610 | 0.4748 | 0.2382 | 0.2016 | 0.2278 | 0.2151 | 0.3325 |
| 4 | 0.2256 | 0.2175 | 0.3681 | 0.2983 | 0.2439 | 0.2504 | 0.2000 | 0.2082 | 0.2014 | 0.2103 |
| 5 | 0.2683 | 0.2115 | 0.3058 | 0.2248 | 0.2234 | 0.2210 | 0.2000 | 0.2080 | 0.2000 | 0.2102 |
| 6 | 0.2000 | 0.2005 | 0.2083 | 0.2229 | 0.2099 | 0.2000 | 0.2000 | 0.2042 | 0.2000 | 0.2039 |
| 7 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 8 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 9 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 10 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
| 11+ | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |
|  | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |  |  |  |
| 3 | 0.4187 | 0.4301 | 0.4276 | 0.2205 | 0.2408 | 0.4019 | 0.4437 |  |  |  |
| 4 | 0.2635 | 0.2798 | 0.3009 | 0.2230 | 0.2202 | 0.2608 | 0.2664 |  |  |  |
| 5 | 0.2081 | 0.2207 | 0.2533 | 0.2178 | 0.2231 | 0.3527 | 0.2674 |  |  |  |
| 6 | 0.2000 | 0.2000 | 0.2204 | 0.2145 | 0.2097 | 0.2546 | 0.2394 |  |  |  |
| 7 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |  |  |  |
| 8 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |  |  |  |
| 9 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |  |  |  |
| 10 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |  |  |  |
| 11+ | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |  |  |  |

Table 4.8. Northeast Arctic haddock. Proportion mature at age

| year | 1950-1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.034 | 0.031 | 0.073 | 0.071 | 0.073 | 0.056 | 0.028 | 0.022 | 0.024 | 0.029 |
| 4 | 0.109 | 0.068 | 0.104 | 0.199 | 0.208 | 0.216 | 0.167 | 0.094 | 0.073 | 0.077 |
| 5 | 0.298 | 0.189 | 0.233 | 0.298 | 0.483 | 0.498 | 0.506 | 0.434 | 0.26 | 0.215 |
| 6 | 0.579 | 0.579 | 0.439 | 0.478 | 0.597 | 0.777 | 0.772 | 0.754 | 0.701 | 0.526 |
| 7 | 0.806 | 0.809 | 0.808 | 0.688 | 0.728 | 0.82 | 0.916 | 0.915 | 0.911 | 0.886 |
| 8 | 0.926 | 0.929 | 0.928 | 0.927 | 0.867 | 0.888 | 0.934 | 0.972 | 0.971 | 0.971 |
| 9 | 0.975 | 0.977 | 0.977 | 0.977 | 0.977 | 0.952 | 0.961 | 0.978 | 0.991 | 0.991 |
| 10 | 0.992 | 0.993 | 0.993 | 0.993 | 0.993 | 0.993 | 0.984 | 0.987 | 0.992 | 0.997 |
| 11+ | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| year | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| 3 | 0.039 | 0.058 | 0.049 | 0.035 | 0.018 | 0.016 | 0.016 | 0.022 | 0.03 | 0.04 |
| 4 | 0.094 | 0.129 | 0.179 | 0.153 | 0.112 | 0.063 | 0.053 | 0.059 | 0.074 | 0.098 |
| 5 | 0.23 | 0.28 | 0.339 | 0.439 | 0.383 | 0.305 | 0.188 | 0.174 | 0.17 | 0.228 |
| 6 | 0.478 | 0.516 | 0.57 | 0.649 | 0.721 | 0.675 | 0.592 | 0.426 | 0.412 | 0.417 |
| 7 | 0.781 | 0.747 | 0.771 | 0.813 | 0.85 | 0.892 | 0.868 | 0.824 | 0.696 | 0.677 |
| 8 | 0.961 | 0.918 | 0.901 | 0.913 | 0.932 | 0.947 | 0.964 | 0.955 | 0.935 | 0.875 |
| 9 | 0.991 | 0.988 | 0.973 | 0.966 | 0.971 | 0.977 | 0.983 | 0.989 | 0.986 | 0.979 |
| 10 | 0.997 | 0.997 | 0.996 | 0.991 | 0.989 | 0.99 | 0.993 | 0.995 | 0.996 | 0.996 |
| 11+ | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| year | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| 3 | 0.055 | 0.035 | 0.035 | 0.024 | 0.025 | 0.029 | 0.033 | 0.028 | 0.013 | 0.009 |
| 4 | 0.127 | 0.165 | 0.109 | 0.113 | 0.085 | 0.08 | 0.092 | 0.104 | 0.088 | 0.05 |
| 5 | 0.286 | 0.34 | 0.403 | 0.302 | 0.307 | 0.233 | 0.235 | 0.261 | 0.296 | 0.261 |
| 6 | 0.509 | 0.575 | 0.637 | 0.697 | 0.599 | 0.594 | 0.509 | 0.518 | 0.549 | 0.585 |
| 7 | 0.684 | 0.752 | 0.817 | 0.855 | 0.884 | 0.821 | 0.824 | 0.765 | 0.776 | 0.797 |
| 8 | 0.862 | 0.864 | 0.905 | 0.933 | 0.948 | 0.961 | 0.939 | 0.939 | 0.912 | 0.914 |
| 9 | 0.956 | 0.951 | 0.953 | 0.969 | 0.978 | 0.983 | 0.987 | 0.98 | 0.979 | 0.97 |
| 10 | 0.993 | 0.986 | 0.984 | 0.984 | 0.989 | 0.993 | 0.995 | 0.996 | 0.994 | 0.994 |
| 11+ | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| year | 2009 |  |  |  |  |  |  |  |  |  |
| 3 | 0.005 |  |  |  |  |  |  |  |  |  |
| 4 | 0.034 |  |  |  |  |  |  |  |  |  |
| 5 | 0.161 |  |  |  |  |  |  |  |  |  |
| 6 | 0.528 |  |  |  |  |  |  |  |  |  |
| 7 | 0.82 |  |  |  |  |  |  |  |  |  |
| 8 | 0.926 |  |  |  |  |  |  |  |  |  |
| 9 | 0.972 |  |  |  |  |  |  |  |  |  |
| 10 | 0.99 |  |  |  |  |  |  |  |  |  |
| 11+ | 1 |  |  |  |  |  |  |  |  |  |

Table 4.9. Northeast Arctic haddock. Survey indices used in tuning XSA

| North-East Arctic haddock |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 103 |  |  |  |  |  |  |  |
| FLT01: Russian BT surve y, total a rea, Nov-Dec, age 1-7 |  |  |  |  |  |  |  |
| 19832009 |  |  |  |  |  |  |  |
| 110.91 .00 |  |  |  |  |  |  |  |
| 17 |  |  |  |  |  |  |  |
| 1 | 592 | 95 | 5 | 4 | 0.1 | 0 | 0 |
| 1 | 586 | 584 | 15 | 2 | 1 | 0.1 | 0 |
| 1 | 144 | 1343 | 900 | 4 | 1 | 1 | 0 |
| 1 | 14 | 107 | 363 | 164 | 1 | 0.1 | 0.1 |
| 1 | 9 | 17 | 83 | 225 | 57 | 0.1 | 0.1 |
| 1 | 3 | 7 | 17 | 40 | 76 | 8 | 0.1 |
| 1 | 18 | 24 | 4 | 14 | 41 | 81 | 11 |
| 1 | -11 | -11 | -11 | -11 | -11 | -11 | -11 |
| 1 | 429 | 176 | 62 | 9 | 3 | 6 | 18 |
| 1 | 282 | 1286 | 346 | 50 | 4 | 6 | 9 |
| 1 | 48 | 357 | 1985 | 356 | 48 | 8 | 4 |
| 1 | 49 | 58 | 442 | 1014 | 116 | 15 | 1 |
| 1 | 72 | 42 | 31 | 123 | 370 | 40 | 5 |
| 1 | 23 | 57 | 28 | 49 | 362 | 334 | 29 |
| 1 | 46 | 19 | 32 | 32 | 10 | 27 | 10 |
| 1 | 29 | 115 | 38 | 46 | 8 | 5 | 15 |
| 1 | 289 | 61 | 196 | 39 | 37 | 8 | 3 |
| 1 | 207 | 262 | 60 | 109 | 26 | 11 | 2 |
| 1 | 149 | 261 | 334 | 40 | 65 | 11 | 4 |
| 1 | 193 | 189 | 399 | 450 | 47 | 24 | 4 |
| 1 | 328 | 251 | 221 | 299 | 231 | 34 | 16 |
| 1 | 110 | 206 | 113 | 94 | 107 | 87 | 5 |
| 1 | 792 | 136 | 240 | 86 | 48 | 57 | 24 |
| 1 | 792 | 1227 | 113 | 119 | 57 | 26 | 24 |
| 1 | 839 | 2142 | 838 | 73 | 137 | 38 | 14 |
| 1 | 127 | 2327 | 2557 | 1051 | 124 | 111 | 17 |
| 1 | 29 | 158 | 1647 | 1704 | 631 | 57 | 32 |

Table 4.9 (continued).

FLT02: Norwegian acoustic, age 1-7, shifted
19802009
110.991 .00

17

| 1 | 140 | 50 | 210 | 600 | 180 | 10 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 20 | 30 | 40 | 40 | 100 | 60 | 0 |
| 1 | 50 | 20 | 30 | 10 | 10 | 40 | 20 |
| 1 | 1730 | 60 | 20 | 10 | 0 | 0 | 0 |
| 1 | 7760 | 2150 | 50 | 0 | 0 | 0 | 0 |
| 1 | 2660 | 4520 | 1890 | 0 | 0 | 0 | 0 |
| 1 | 170 | 490 | 1710 | 500 | 0 | 0 | 0 |
| 1 | 40 | 80 | 230 | 460 | 70 | 0 | 0 |
| 1 | 50 | 60 | 110 | 200 | 210 | 20 | 0 |
| 1 | 350 | 30 | 30 | 40 | 70 | 110 | 20 |
| 1 | 2520 | 450 | 80 | 30 | 30 | 30 | 60 |
| 1 | 8680 | 1340 | 230 | 20 | 0 | 0 | 10 |
| 1 | 6260 | 5630 | 1300 | 130 | 0 | 0 | 0 |
| 1 | 1930 | 2550 | 6310 | 1110 | 120 | 0 | 0 |
| 1 | 2850 | 360 | 1110 | 3870 | 420 | 20 | 0 |
| 1 | 2290 | 440 | 310 | 760 | 1510 | 80 | 0 |
| 1 | 240 | 510 | 170 | 120 | 430 | 430 | 20 |
| 1 | 1220 | 200 | 280 | 120 | 50 | 130 | 160 |
| 1 | 460 | 570 | 130 | 140 | 40 | 10 | 20 |
| 1 | 5090 | 320 | 650 | 190 | 110 | 20 | 10 |
| 1 | 3160 | 2100 | 230 | 220 | 10 | 10 | 0 |
| 1 | 2820 | 2160 | 1490 | 140 | 120 | 10 | 0 |
| 1 | 2790 | 1450 | 1980 | 1690 | 170 | 50 | 0 |
| 1 | 4740 | 1270 | 760 | 760 | 660 | 70 | 20 |
| 1 | 2090 | 2190 | 1020 | 360 | 400 | 90 | 0 |
| 1 | 8040 | 540 | 860 | 300 | 120 | 90 | 20 |
| 1 | 8680 | 3790 | 540 | 880 | 220 | 60 | 50 |
| 1 | 18352 | 7234 | 2517 | 573 | 742 | 102 | 58 |
| 1 | 2463 | 10217 | 7730 | 4021 | 313 | 149 | 16 |
| 1 | 818 | 1380 | 5930 | 5574 | 1914 | 103 | 29 |

Table 4.9 (continued).

| FLT04: Norwe gian BT survey, age 1-8, shifted |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19822009 |  |  |  |  |  |  |  |  |
| 110.991 .00 |  |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |  |
| 1 | 48 | 31 | 24 | 9 | 19 | 25 | 7 | 0 |
| 1 | 5146 | 189 | 15 | 8 | 2 | 1 | 4 | 1 |
| 1 | 15938 | 4759 | 147 | 5 | 5 | 1 | 1 | 4 |
| 1 | 3703 | 3846 | 1108 | 6 | 2 | 1 | 1 | 1 |
| 1 | 799 | 1544 | 2902 | 529 | 0 | 0 | 0 | 0 |
| 1 | 153 | 253 | 689 | 1164 | 138 | 1 | 0 | 0 |
| 1 | 95 | 141 | 216 | 340 | 327 | 34 | 1 | 0 |
| 1 | 546 | 45 | 34 | 50 | 92 | 118 | 18 | 0 |
| 1 | 3003 | 334 | 51 | 42 | 27 | 17 | 42 | 0 |
| 1 | 13755 | 1505 | 244 | 21 | 6 | 7 | 16 | 23 |
| 1 | 5990 | 5077 | 1056 | 105 | 6 | 4 | 3 | 4 |
| 1 | 2280 | 3395 | 4366 | 497 | 34 | 2 | 1 | 2 |
| 1 | 1793 | 536 | 1711 | 3395 | 345 | 28 | 0 | 1 |
| 1 | 2636 | 525 | 481 | 1486 | 2528 | 116 | 9 | 0 |
| 1 | 679 | 861 | 280 | 194 | 467 | 622 | 35 | 1 |
| 1 | 1379 | 227 | 332 | 132 | 34 | 80 | 81 | 7 |
| 1 | 576 | 598 | 122 | 102 | 28 | 10 | 17 | 11 |
| 1 | 4522 | 272 | 354 | 84 | 40 | 8 | 3 | 7 |
| 1 | 4603 | 2960 | 293 | 251 | 17 | 9 | 1 | 1 |
| 1 | 5347 | 3147 | 1853 | 176 | 82 | 8 | 3 | 0 |
| 1 | 5131 | 3174 | 1820 | 736 | 55 | 23 | 2 | 1 |
| 1 | 7112 | 1881 | 1027 | 804 | 462 | 59 | 11 | 2 |
| 1 | 4204 | 3465 | 1333 | 668 | 522 | 123 | 6 | 2 |
| 1 | 13131 | 774 | 1405 | 482 | 196 | 152 | 31 | 1 |
| 1 | 15938 | 5077 | 660 | 860 | 233 | 75 | 37 | 14 |
| 1 | 21294 | 15224 | 6009 | 868 | 489 | 62.7 | 25.1 | 8.2 |
| 1 | 3280 | 12704 | 7732 | 3654 | 385 | 106 | 14 | 1 |
| 1 | 1112 | 1028 | 5086 | 4796 | 1312 | 70 | 10 | 6 |

Table 4.9A. Northeast Arctic haddock. Ecosystem survey indices used in tuning XSA

North-East Arctic haddock
101
FLT007: Ecosystem, total area, avg-sep, age 1-8
\#\#\#\#\#\#\#
110.650 .75

18

| 1 | 189 | 269 | 123 | 70 | 69 | 31 | 3 | 2 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 627 | 114 | 323 | 89 | 29 | 31 | 15 | 0 |
| 1 | 2152 | 1068 | 151 | 143 | 43 | 18 | 16 | 6 |
| 1 | 995 | 1827 | 1287 | 89 | 94 | 19 | 6 | 7 |
| 1 | 322 | 1293 | 1155 | 406 | 43 | 36 | 5 | 3 |
| 1 | 136 | 143 | 649 | 617 | 306 | 21 | 7 | 1 |

Table 4.10. Northeast Arctic haddock. Input data for recruitment prediction (RCT3)
NORTHEAST ARCTIC HADDOCK: recruits as 3 year-olds 9202
'Year-

| class' |  | 'VPA' | 'NT1' | 'NT2' | 'NT3' | 'NAK1' | 'NAK2' | 'NAK3' | 'RT1' | 'RT2' | 'RT3' |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1990 | 687 | 2006 | 1375.5 | 507.7 | 1890 | 868 | 563 | -11 | 42.9 | 128.6 |
|  | 1991 | 303 | 1659.4 | 599 | 339.5 | 1135 | 626 | 255 | 16.7 | 28.2 | 35.7 |
|  | 1992 | 99 | 727.9 | 228 | 53.6 | 947 | 193 | 36 | 16.4 | 4.8 | 5.8 |
|  | 1993 | 105 | 603.2 | 179.3 | 52.5 | 562 | 285 | 44 | 3.5 | 4.9 | 4.2 |
|  | 1994 | 119 | 1463.6 | 263.6 | 86.1 | 1379 | 229 | 51 | 9.1 | 7.2 | 5.7 |
|  | 1995 | 59 | 309.5 | 67.9 | 22.7 | 249 | 24 | 20 | 6.4 | 2.3 | 1.9 |
|  | 1996 | 231 | 1268 | 137.9 | 59.8 | 693 | 122 | 57 | 6 | 4.6 | 11.5 |
|  | 1997 | 85 | 212.9 | 57.6 | 27.2 | 220 | 46 | 32 | 1.8 | 2.9 | 6.1 |
|  | 1998 | 370 | 1244.9 | 452.2 | 296 | 856 | 509 | 210 | 10.7 | 28.9 | 26.2 |
|  | 1999 | 342 | 847.2 | 460.3 | 314.7 | 1024 | 316 | 216 | 11.7 | 20.7 | 26.1 |
|  | 2000 | 222 | 1220.5 | 534.7 | 317.4 | 976 | 282 | 145 | 15.1 | 14.9 | 18.9 |
|  | 2001 | 237 | 1680.3 | 513.1 | 188.1 | 2062 | 279 | 127 | 20.8 | 19.3 | 25.1 |
|  | 2002 | 371 | 3332.1 | 711.2 | 346.5 | 2394 | 474 | 219 | 33.2 | 32.8 | 20.6 |
|  | 2003 | 185 | 715.9 | 420.4 | 77.4 | 752 | 209 | 54 | 19.8 | 11 | 13.6 |
|  | 2004 | 610 | 4630.2 | 1313.1 | 507.7 | 3364 | 804 | 379 | 50 | 79.2 | 122.7 |
|  | 2005 | 1029 | 5141.3 | 1593.8 | 1522.4 | 2767 | 868 | 723.4 | 62 | 79.2 | 214.2 |
|  | 2006 | 811 | 3874.4 | 2129.4 | 1270 | 3197 | 1835.2 | 1021.7 | 53.4 | 83.9 | 232.7 |
|  | 2007 | -11 | 860.2 | 328 | 102.8 | 1266.6 | 246.3 | 138 | 6.5 | 12.7 | 15.8 |
|  | 2008 | -11 | 564.7 | 111.2 | -11 | 849 | 81.8 | -11 | 5.7 | 2.9 | -11 |
|  | 2009 | -11 | 1619.5 | -11 | -11 | 2035.8 | -11 | -11 | 10 | -11 | -11 |

1990 RT was removed from XSA tuning
RT1 Russian bottom trawl survey age 1
RT 2 Russian bottom trawl survey age 2
RT 3 Russian bottom trawl survey age 3
NT 1 Norwegian bottom trawl survey age 1
NT 2 Norwegian bottom trawl survey age 2
NT 3 Norwegian bottom trawl survey age 3
NA1 Norwegian acoustic survey age 1
NA2 Norwegian acoustic survey age 2
NA3 Norwegian acoustic survey age 3

Table 4.11. Northeast Arctic haddock. Analysis by RCT3 ver. 1

Data for 9 surveys over 20 years: 1990-2009 Regression type = C Tapered time weighting applied power $=3$ over 20 years Survey weighting not applied Final estimates shrunk towards mean Minimum S.E. for any survey taken as 0.20 Minimum of 3 points used for regression
Forecast/Hindcast variance correction used.

| Yearclass | $=$ | 2004 | Regression |  |  |  | Prediction |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey/ | Slope | Inter- | Std | Rsquare | No. | Index | Predicted | Std | WAP |
| Series |  | cept | Error |  | Pts | Value | Value | Error | Weights |
| NT1 | 1.11 | -2.34 | 0.56 | 0.597 | 14 | 8.44 | 7 | 0.752 | 0.038 |
| NT2 | 0.87 | 0.3 | 0.42 | 0.723 | 14 | 7.18 | 6.56 | 0.541 | 0.073 |
| NT3 | 0.7 | 1.91 | 0.31 | 0.826 | 14 | 6.23 | 6.26 | 0.388 | 0.143 |
| NAK1 | 1.23 | -3.05 | 0.64 | 0.53 | 14 | 8.12 | 6.94 | 0.842 | 0.03 |
| NAK2 | 0.82 | 0.85 | 0.45 | 0.699 | 14 | 6.69 | 6.36 | 0.556 | 0.069 |
| NAK3 | 0.75 | 1.87 | 0.23 | 0.901 | 14 | 5.94 | 6.35 | 0.285 | 0.263 |
| RT1 | 1.38 | 1.82 | 0.81 | 0.388 | 13 | 3.93 | 7.24 | 1.096 | 0.018 |
| RT2 | 0.84 | 3.18 | 0.29 | 0.843 | 14 | 4.38 | 6.85 | 0.405 | 0.131 |
| RT3 | 0.8 | 3.15 | 0.24 | 0.89 | 14 | 4.82 | 6.99 | 0.341 | 0.184 |
| VPA | Mean | = |  |  |  |  | 5.28 | 0.647 | 0.051 |
| Yearclass | $=$ | 2005 | Regression |  |  |  | Prediction |  |  |
| Survey/ | Slope | Inter- | Std | Rsquare | No. | Index | Predicted | Std | WAP |
| Series |  | cept | Error |  | Pts | Value | V alue | Error | Weights |
| NT1 | 0.97 | -1.45 | 0.49 | 0.688 | 15 | 8.55 | 6.87 | 0.632 | 0.051 |
| NT2 | 0.84 | 0.46 | 0.39 | 0.775 | 15 | 7.37 | 6.67 | 0.497 | 0.083 |
| NT3 | 0.71 | 1.84 | 0.31 | 0.846 | 15 | 7.33 | 7.08 | 0.425 | 0.113 |
| NAK1 | 1.09 | -2.14 | 0.56 | 0.63 | 15 | 7.93 | 6.51 | 0.681 | 0.044 |
| NAK2 | 0.82 | 0.86 | 0.42 | 0.752 | 15 | 6.77 | 6.44 | 0.512 | 0.078 |
| NAK3 | 0.77 | 1.83 | 0.22 | 0.917 | 15 | 6.59 | 6.87 | 0.292 | 0.239 |
| RT1 | 1.15 | 2.34 | 0.66 | 0.532 | 14 | 4.14 | 7.11 | 0.877 | 0.027 |
| RT2 | 0.77 | 3.33 | 0.28 | 0.873 | 15 | 4.38 | 6.69 | 0.357 | 0.161 |
| RT3 | 0.72 | 3.31 | 0.25 | 0.893 | 15 | 5.37 | 7.2 | 0.355 | 0.162 |
| VPA | Mean | = |  |  |  |  | 5.38 | 0.691 | 0.043 |
| Yearclass | $=$ | 2006 | Regression |  |  |  | Prediction |  |  |
| Survey/ | Slope | Inter- | Std | Rsquare | No. | Index | Predicted | Std | WAP |
| Series |  | cept | Error |  | Pts | V alue | V alue | Error | Weights |
| NT1 | 0.97 | -1.42 | 0.47 | 0.761 | 16 | 8.26 | 6.59 | 0.561 | 0.059 |
| NT2 | 0.89 | 0.22 | 0.4 | 0.813 | 16 | 7.66 | 7.01 | 0.507 | 0.072 |
| NT3 | 0.69 | 1.95 | 0.29 | 0.892 | 16 | 7.15 | 6.89 | 0.365 | 0.14 |
| NAK1 | 1.17 | -2.67 | 0.58 | 0.671 | 16 | 8.07 | 6.79 | 0.71 | 0.037 |
| NAK2 | 0.91 | 0.44 | 0.45 | 0.775 | 16 | 7.52 | 7.27 | 0.585 | 0.054 |
| NAK3 | 0.78 | 1.78 | 0.21 | 0.939 | 16 | 6.93 | 7.17 | 0.28 | 0.238 |
| RT1 | 1.1 | 2.47 | 0.6 | 0.653 | 15 | 4 | 6.85 | 0.743 | 0.034 |
| RT2 | 0.8 | 3.25 | 0.28 | 0.896 | 16 | 4.44 | 6.81 | 0.354 | 0.149 |
| RT3 | 0.69 | 3.4 | 0.24 | 0.923 | 16 | 5.45 | 7.15 | 0.315 | 0.188 |
| VPA | Mean | $=$ |  |  |  |  | 5.53 | 0.79 | 0.03 |

Table 4.11 (continued).

| Yearclass | $=$ | 2007 | Regression |  |  |  | Prediction |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey/ | Slope | Inter- | Std | Rsquare | No. | Index | Predicted | Std | WAP |
| Series |  | cept | Error |  | Pts | V alue | V alue | Error | Weights |
| NT1 | 0.97 | -1.4 | 0.44 | 0.791 | 17 | 6.76 | 5.15 | 0.511 | 0.055 |
| NT2 | 0.85 | 0.44 | 0.37 | 0.844 | 17 | 5.8 | 5.35 | 0.425 | 0.08 |
| NT3 | 0.67 | 2.05 | 0.27 | 0.91 | 17 | 4.64 | 5.16 | 0.313 | 0.147 |
| NAK1 | 1.14 | -2.47 | 0.54 | 0.719 | 17 | 7.14 | 5.7 | 0.614 | 0.038 |
| NAK2 | 0.84 | 0.82 | 0.4 | 0.819 | 17 | 5.51 | 5.43 | 0.463 | 0.067 |
| NAK3 | 0.73 | 1.99 | 0.22 | 0.936 | 17 | 4.93 | 5.58 | 0.256 | 0.22 |
| RT1 | 1.06 | 2.56 | 0.55 | 0.71 | 16 | 2.01 | 4.7 | 0.652 | 0.034 |
| RT2 | 0.79 | 3.28 | 0.26 | 0.914 | 17 | 2.62 | 5.34 | 0.303 | 0.156 |
| RT3 | 0.65 | 3.51 | 0.25 | 0.924 | 17 | 2.82 | 5.33 | 0.283 | 0.18 |
| VPA | Mean | $=$ |  |  |  |  | 5.65 | 0.82 | 0.021 |
| Yearclass | $=$ | 2008 | Regression |  |  |  | Prediction |  |  |
| Survey/ | Slope | Inter- | Std | Rsquare | No. | Index | Predicted | Std | WAP |
| Series |  | cept | Error |  | Pts | V alue | V alue | Error | Weights |
| NT1 | 0.95 | -1.28 | 0.43 | 0.798 | 17 | 6.34 | 4.76 | 0.523 | 0.13 |
| NT2 | 0.84 | 0.45 | 0.37 | 0.845 | 17 | 4.72 | 4.44 | 0.461 | 0.167 |
| NT3 |  |  |  |  |  |  |  |  |  |
| NAK1 | 1.13 | -2.37 | 0.53 | 0.726 | 17 | 6.75 | 5.25 | 0.616 | 0.094 |
| NAK2 | 0.83 | 0.87 | 0.39 | 0.831 | 17 | 4.42 | 4.53 | 0.481 | 0.154 |
| NAK3 |  |  |  |  |  |  |  |  |  |
| RT1 | 1.05 | 2.59 | 0.54 | 0.72 | 16 | 1.9 | 4.58 | 0.656 | 0.083 |
| RT2 | 0.78 | 3.29 | 0.26 | 0.915 | 17 | 1.36 | 4.35 | 0.334 | 0.319 |
| RT3 |  |  |  |  |  |  |  |  |  |
| VPA | Mean | $=$ |  |  |  |  | 5.69 | 0.816 | 0.053 |
| Yearclass | $=$ | 2009 | Regression |  |  |  | Prediction |  |  |
| Survey/ | Slope | Inter- | Std | Rsquare | No. | Index | Predicted | Std | WAP |
| Series |  | cept | Error |  | Pts | V alue | Value | Error | Weights |
| NT1 | 0.94 | -1.16 | 0.42 | 0.804 | 17 | 7.39 | 5.76 | 0.495 | 0.377 |
| NT2 |  |  |  |  |  |  |  |  |  |
| NT3 |  |  |  |  |  |  |  |  |  |
| NAK1 | 1.12 | $-2.28$ | 0.52 | 0.732 | 17 | 7.62 | 6.23 | 0.615 | 0.244 |
| NAK2 |  |  |  |  |  |  |  |  |  |
| NAK3 |  |  |  |  |  |  |  |  |  |
| RT1 | 1.04 | 2.61 | 0.52 | 0.73 | 16 | 2.4 | 5.1 | 0.625 | 0.237 |
| RT2 RT3 |  |  |  |  |  |  |  |  |  |
| VPA | Mean | = |  |  |  |  | 5.74 | 0.808 | 0.142 |
| Year | Weighted | Log | Int | Ext | V ar | VPA | Log |  |  |
| Class | Average | WAP | Std | Std |  | Ratio | VPA |  |  |
| Prediction |  |  | Error | Error |  |  |  |  |  |
| 2004 | 691 | 6.54 | 0.15 | 0.14 | 0.91 | 610 | 6.42 |  |  |
| 2005 | 892 | 6.79 | 0.14 | 0.13 | 0.77 | 1029 | 6.94 |  |  |
| 2006 | 1052 | 6.96 | 0.14 | 0.1 | 0.58 | 812 | 6.7 |  |  |
| 2007 | 212 | 5.36 | 0.12 | 0.07 | 0.32 |  |  |  |  |
| 2008 | 101 | 4.62 | 0.19 | 0.15 | 0.61 |  |  |  |  |
| 2009 | 303 | 5.71 | 0.3 | 0.23 | 0.56 |  |  |  |  |

Table 4.12. Northeast Arctic haddock. Extended Survivors Analysis
FLR XSA Diagnostics 2010-04-23 18:00:06
CPUE data from indices

Catch data for 59 years. 1950 to 2009. Ages 1 to 11.

## fleet

1 FLT01: Russian BT survey, totalarea, Nov-Dec, age 1-7
2 FLT02: Norwe gian a coustic, age 1-7, shifted
3 FLT04: Norwe gian BT survey, age $1-8$, shifted
Time se ries weights :
Ta pered time weighting applied Power = 3 over 20 years
Catchability a nalysis :
Catchability inde pendent of size for ages $>6$
Catchability inde pendent of age forages $>9$
Terminal populationestimation:
Survivor estimates shrunk towards the mean F
of the final 5 years or the 3 oldestages.
S.E of the mean to which the estimates a re shrunk $=0.5$

Minimum standarderror for population
estimates de rived from each flee $t=0.3$
prior weighting not applied
Regression weights

|  | year |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| all | 0.751 | 0.82 | 0.877 | 0.921 | 0.954 | 0.976 | 0.99 | 0.997 | 1 |
|  | Fishing year | mortalities |  |  |  |  |  |  |  |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| 3 | 0.02 | 0.04 | 0.025 | 0.039 | 0.043 | 0.039 | 0.028 | 0.058 | 0.017 |
| 4 | 0.213 | 0.091 | 0.198 | 0.162 | 0.19 | 0.165 | 0.184 | 0.125 | 0.118 |
| 5 | 0.279 | 0.374 | 0.286 | 0.398 | 0.35 | 0.351 | 0.402 | 0.404 | 0.205 |
| 6 | 0.274 | 0.386 | 0.478 | 0.563 | 0.575 | 0.632 | 0.429 | 0.53 | 0.559 |
| 7 | 0.349 | 0.249 | 0.334 | 0.686 | 0.493 | 0.828 | 0.634 | 0.628 | 0.55 |
| 8 | 0.357 | 0.276 | 0.208 | 0.363 | 0.522 | 0.473 | 0.802 | 0.761 | 0.604 |
| 9 | 0.267 | 0.457 | 0.18 | 0.485 | 0.217 | 0.716 | 0.492 | 0.599 | 0.697 |
| 10 | 0.322 | 0.326 | 0.319 | 0.559 | 0.577 | 0.676 | 0.806 | 0.541 | 0.617 |
| 11 | 0.322 | 0.326 | 0.319 | 0.559 | 0.577 | 0.676 | 0.806 | 0.541 | 0.617 |
|  | XSA | population | number | (Thousand) |  |  |  |  |  |
|  | age |  |  |  |  |  |  |  |  |
| year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
| 2000 | 84937 | 173204 | 29624 | 23799 | 9026 | 6105 | 7405 | 6997 | 1741 |
| 2001 | 370480 | 66278 | 113634 | 18206 | 14754 | 5213 | 3499 | 4642 | 4748 |
| 2002 | 342198 | 287138 | 49462 | 64036 | 10129 | 9418 | 3239 | 1814 | 4138 |
| 2003 | 221794 | 239348 | 190968 | 30109 | 32370 | 5939 | 6262 | 2216 | 3114 |
| 2004 | 237499 | 140410 | 156341 | 104191 | 14032 | 13345 | 3381 | 3156 | 2246 |
| 2005 | 370823 | 148062 | 87786 | 88385 | 48002 | 7019 | 6480 | 2227 | 3385 |
| 2006 | 185450 | 232575 | 92926 | 47979 | 37702 | 17167 | 3582 | 2593 | 1413 |
| 2007 | 609862 | 144690 | 154738 | 50012 | 25211 | 16370 | 6304 | 1793 | 2585 |
| 2008 | 1028683 | 452155 | 102469 | 82658 | 23877 | 11012 | 6264 | 2835 | 1552 |
| 2009 | 810831 | 676498 | 309587 | 58672 | 36635 | 11274 | 4926 | 2553 | 2737 |


|  | Estimated <br> age | population | abundance | at | 1st | Jan | 2009 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 3 |  |  |  |  |  |  |  |  |
| year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
| 2009 | 0 | 508063 | 470158 | 191039 | 34048 | 16076 | 5793 | 2523 | 1288 |

Table 4.12 (continued).
Fleet: $\quad 1$ FLT01: Russian BT survey, total area, Nov-Dec, age 1-7
Log catchability residuals.

| year | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age |  |  |  |  |  |  |  |  |  |  |  |
| 3 | NA | 0.037 | 0.265 | 0.201 | 0.148 | -0.143 | -0.067 | -0.195 | 0.215 | 0.002 | 0.156 |
| 4 | NA | -0.003 | -0.019 | 0.399 | 0.066 | -0.246 | 0.114 | 0.151 | 0.064 | 0.29 | -0.15 |
| 5 | NA | -0.249 | -0.251 | 0.147 | 0.129 | -0.204 | 0.45 | -0.413 | -0.293 | 0.198 | 0.259 |
| 6 | NA | -0.3 | 0.279 | 0.382 | -0.013 | -0.012 | 0.161 | -0.381 | -0.479 | -0.004 | -0.148 |
| 7 | NA | 0.104 | 0.305 | 0.507 | -0.793 | -0.013 | 0.973 | -1.334 | 0.001 | -0.638 | -0.877 |
| age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |  |  |
| 3 | -0.1 | 0.052 | 0.131 | -0.157 | -0.149 | -0.085 | -0.045 | 0.096 | 0.094 |  |  |
| 4 | -0.007 | 0.132 | 0.081 | -0.056 | -0.132 | -0.313 | -0.239 | 0.192 | 0.131 |  |  |
| 5 | -0.193 | 0.188 | 0.051 | -0.182 | -0.144 | -0.097 | -0.037 | 0.167 | 0.132 |  |  |
| 6 | 0.098 | -0.356 | 0.408 | 0.007 | -0.056 | -0.126 | 0.088 | 0.318 | 0.087 |  |  |
| 7 | -0.77 | -0.314 | 0.245 | -0.266 | 0.392 | 0.449 | 0.307 | 0.481 | 0.755 |  |  |

Mean $\log$ catchability and standard error of ages with catchability
independent of year class streng th and constant w.r.t. time
Mean_Logq S.E_Logq
-7.016 0.622
Regression statistics
Ages with q dependent on year class streng th

|  |  | slope | intercept |
| :--- | :--- | :--- | :--- |
| Age | 3 | 0.6123 | 8.952543 |
| Age | 4 | 0.6628 | 8.461045 |
| Age | 5 | 0.6778 | 8.157107 |
| Age | 6 | 0.7197 | 7.819472 |

Table 4.12 (continued).
Fleet: $\quad 2$ FLT02: Norwegian acoustic, age 1-7, shifted
Log catchability residuals.

| year <br> age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3 | 0.166 | -0.11 | 0.25 | 0.195 | -0.059 | 0.132 | 0.004 | 0.021 | 0.026 | -0.138 | 0.038 |
| 4 | 0.094 | -0.255 | -0.197 | 0.307 | 0.064 | -0.038 | -0.09 | 0.135 | -0.043 | 0.392 | -0.439 |
| 5 | 0.036 | NA | NA | 0.116 | 0.17 | -0.137 | 0.01 | -0.04 | 0.075 | 0.229 | -0.373 |
| 6 | -0.275 | NA | NA | NA | -0.113 | 0.059 | -0.027 | 0.133 | -0.345 | 0.207 | -0.42 |
| 7 | 0.405 | -1.26 | NA | NA | NA | NA | -0.152 | 0.686 | -0.472 | -0.197 | NA |


| age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | -0.083 | 0.157 | 0.046 | 0.147 | -0.247 | -0.094 | -0.156 | 0.077 | 0.138 |
| 4 | -0.056 | 0.123 | -0.102 | -0.059 | -0.177 | -0.026 | 0.072 | 0.193 | 0.063 |
| 5 | -0.292 | 0.279 | -0.013 | -0.076 | -0.133 | 0.041 | 0.155 | 0.082 | 0.023 |
| 6 | -0.184 | -0.24 | 0.479 | -0.252 | -0.102 | 0.028 | 0.297 | 0.166 | 0.099 |
| 7 | NA | NA | -0.287 | NA | -0.539 | 0.425 | 0.97 | -0.341 | -0.101 |

Mean $\log$ catchability and standard error of ages with catchability
independent of year class streng th and constant w.r.t. time
Mean_Logq S.E_Logq
-6.221 0.608
Regression statistics
Ages with q dependent on year class streng th

|  |  | slope | intercept |
| :--- | :--- | :--- | :--- |
| Age | 3 | 0.6922 | 7.4686 |
| Age | 4 | 0.6663 | 7.5303 |
| Age | 5 | 0.5775 | 8.0407 |
| Age | 6 | 0.6899 | 7.5522 |

Table 4.12 (continued).

Fleet: $\quad 3$ FLT04: Norwegian BT, age 1-8, shifted
Log catchability residuals.

| year | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age |  |  |  |  |  |  |  |  |  |  |  |
| 3 | -0.14 | -0.171 | 0.063 | -0.067 | 0.062 | 0.255 | 0.155 | 0.014 | -0.094 | -0.518 | 0.065 |
| 4 | 0.19 | -0.317 | -0.349 | -0.07 | 0.021 | 0.273 | 0.094 | 0.142 | -0.236 | -0.018 | -0.425 |
| 5 | 0.097 | 0.001 | -0.111 | -0.2 | 0.141 | 0.041 | 0.074 | -0.072 | 0.049 | -0.019 | -0.083 |
| 6 | -0.313 | -0.171 | 0.099 | -0.185 | 0.128 | 0.211 | 0.032 | -0.059 | -0.132 | -0.02 | -0.219 |
| 7 | 0.842 | 0.003 | -0.772 | -0.851 | NA | 0.615 | 1.201 | 0.799 | 0.159 | -0.608 | -1.548 |
| 8 | NA | 1.158 | -0.478 | -0.163 | 0.308 | NA | 0.017 | 1.028 | 0.238 | 0.519 | -0.676 |
| age | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |  |  |
| 3 | -0.066 | 0.03 | 0.103 | 0.188 | -0.098 | -0.086 | 0.176 | -0.008 | -0.022 |  |  |
| 4 | 0.009 | -0.266 | -0.089 | 0.212 | 0.018 | -0.045 | 0.258 | 0.174 | 0.018 |  |  |
| 5 | -0.323 | -0.012 | -0.085 | 0.049 | 0.083 | 0.108 | 0.052 | 0.185 | -0.076 |  |  |
| 6 | -0.066 | -0.408 | 0.384 | -0.087 | 0.109 | 0.153 | 0.1 | 0.011 | -0.022 |  |  |
| 7 | -1.04 | -0.985 | -0.092 | -0.054 | 0.692 | 0.917 | 0.926 | 0.319 | -0.373 |  |  |
| 8 | NA | -1.257 | 0.052 | -0.6 | -0.7 | 1.372 | 0.844 | -1.019 | 0.611 |  |  |

Mean $\log$ catchability and standard error of ages with catchability
independent of year class streng th and constant w.r.t. time

$$
7 \quad 8
$$

Mean_Logq $\quad-7.014 \quad-7.487$

| S.E_Logq | $0.808 \quad 0.786$ |
| :--- | :--- | :--- |

Regression statistics
Ages with q dependent on year class streng th

|  |  | slope | intercept |
| :--- | :--- | :--- | :--- |
| Age | 3 | 0.6935 | 7.3331 |
| Age | 4 | 0.6920 | 7.3349 |
| Age | 5 | 0.5436 | 8.3400 |
| Age | 6 | 0.5853 | 8.1385 |

## Table 4.12 (continued).

| Age 3 Year class =2006 | scaledWts | survivors | y rels |
| :---: | :---: | :---: | :---: |
| 1 FLT01: Russian BT survey, total area, Nov-Dec, age 1-7 | 0.285 | 592494 | 2006 |
| 2 FLT02: Norwegian acoustic, age 1-7, shifted | 0.285 | 619827 | 2006 |
| 3 FLT04: Norwegian BT survey, age 1-8, shifted | 0.285 | 492398 | 2006 |
| fshk | 0.105 | 323374 | 2006 |
| nshk | 0.04 | 170846 | 2006 |
| Age 4 Year class $=2005$ | scaledWts | survivors | y rels |
| 1 FLT01: Russian BT survey, total area, Nov-Dec, age 1-7 | 0.281 | 572721 | 2005 |
| 2 FLT02: Norwegian acoustic, age 1-7, shifted | 0.287 | 516596 | 2005 |
| 3 FLT04: Norwegian BT survey, age 1-8, shifted | 0.286 | 482469 | 2005 |
| fshk | 0.147 | 282540 | 2005 |
| Age 5 Year class =2004 | scaledWts | survivors | y rels |
| 1 FLT01: Russian BT survey, total area, Nov-Dec, age 1-7 | 0.234 | 232098 | 2004 |
| 2 FLT02: Norwegian acoustic, age 1-7, shifted | 0.241 | 198633 | 2004 |
| 3 FLT04: Norwegian BT survey, age 1-8, shifted | 0.363 | 166256 | 2004 |
| fshk | 0.162 | 111366 | 2004 |
| Age 6 Year class =2003 | scaledWts | survivors | y rcls |
| 1 FLT01: Russian BT survey, total area, Nov-Dec, age 1-7 | 0.274 | 38439 | 2003 |
| 2 FLT02: Norweg ian acoustic, age 1-7, shifted | 0.223 | 39304 | 2003 |
| 3 FLT04: Norwegian BT survey, age 1-8, shifted | 0.296 | 32783 | 2003 |
| fshk | 0.207 | 16535 | 2003 |
| Age 7 Year class =2002 | scaledWts | survivors | y rels |
| 1 FLT01: Russian BT survey, total area, Nov-Dec, age 1-7 | 0.171 | 34216 | 2002 |
| 2 FLT02: Norwegian acoustic, age 1-7, shifted | 0.214 | 14531 | 2002 |
| 3 FLT04: Norwegian BT survey, age 1-8, shifted | 0.096 | 11074 | 2002 |
| fshk | 0.519 | 15778 | 2002 |
| Age 8 Year class =2001 | scaledWts | survivors | y rels |
| 3 FLT04: Norwegian BT survey, age 1-8, shifted | 0.164 | 10675 | 2001 |
| fshk | 0.836 | 3851 | 2001 |
| Age 9 Year class =2000 | scaledWts | survivors | y rels |
| fshk | 1 | 2066 | 2000 |
| Age 10 Year class = 1999 | scaledWts | survivors | y rels |
| fshk | 1 | 1166 | 1999 |

Table 4.13. Northeast Arctic haddock. Fishing mortalit y at age

|  | 1950 | 1951 | 1952 | 1953 | 1954 | 1955 | 1956 | 1957 | 1958 | 1959 | 1960 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.049 | 0.126 | 0.105 | 0.064 | 0.055 | 0.023 | 0.102 | 0.040 | 0.025 | 0.064 | 0.182 |
| 4 | 0.576 | 0.211 | 0.532 | 0.379 | 0.236 | 0.130 | 0.168 | 0.241 | 0.168 | 0.168 | 0.366 |
| 5 | 0.814 | 0.625 | 0.577 | 0.528 | 0.303 | 0.482 | 0.273 | 0.368 | 0.570 | 0.331 | 0.511 |
| 6 | 0.809 | 0.911 | 0.888 | 0.488 | 0.410 | 0.467 | 0.810 | 0.403 | 0.519 | 0.555 | 0.649 |
| 7 | 1.158 | 0.803 | 0.997 | 0.714 | 0.614 | 1.015 | 0.625 | 0.816 | 0.966 | 0.601 | 0.518 |
| 8 | 1.002 | 1.002 | 1.256 | 0.655 | 0.864 | 0.622 | 0.936 | 0.450 | 0.870 | 0.429 | 0.701 |
| 9 | 0.647 | 1.428 | 1.378 | 0.513 | 1.366 | 0.429 | 0.397 | 0.628 | 0.744 | 0.845 | 1.150 |
| 10 | 0.946 | 1.090 | 1.225 | 0.633 | 0.958 | 0.695 | 0.659 | 0.637 | 0.869 | 0.630 | 0.798 |
| 11+ | 0.946 | 1.090 | 1.225 | 0.633 | 0.958 | 0.695 | 0.659 | 0.637 | 0.869 | 0.630 | 0.798 |
| FBAR4-7 | 0.839 | 0.637 | 0.749 | 0.527 | 0.391 | 0.523 | 0.469 | 0.457 | 0.556 | 0.414 | 0.511 |
|  | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 |
| 3 | 0.154 | 0.181 | 0.109 | 0.073 | 0.060 | 0.117 | 0.055 | 0.037 | 0.090 | 0.153 | 0.021 |
| 4 | 0.472 | 0.579 | 0.660 | 0.308 | 0.231 | 0.375 | 0.298 | 0.384 | 0.163 | 0.225 | 0.259 |
| 5 | 0.686 | 1.052 | 0.928 | 0.684 | 0.461 | 0.589 | 0.416 | 0.572 | 0.491 | 0.243 | 0.177 |
| 6 | 0.749 | 1.060 | 1.027 | 0.871 | 0.697 | 0.743 | 0.519 | 0.458 | 0.581 | 0.502 | 0.180 |
| 7 | 0.832 | 0.698 | 1.002 | 0.846 | 0.677 | 0.826 | 0.533 | 0.704 | 0.405 | 0.530 | 0.402 |
| 8 | 0.880 | 0.901 | 0.649 | 0.961 | 0.596 | 0.528 | 0.581 | 0.718 | 0.503 | 0.413 | 0.389 |
| 9 | 0.964 | 1.183 | 1.362 | 1.389 | 1.053 | 0.593 | 0.383 | 0.495 | 0.502 | 0.395 | 0.296 |
| 10 | 0.902 | 0.937 | 1.016 | 1.078 | 0.783 | 0.655 | 0.503 | 0.645 | 0.473 | 0.449 | 0.365 |
| 11+ | 0.902 | 0.937 | 1.016 | 1.078 | 0.783 | 0.655 | 0.503 | 0.645 | 0.473 | 0.449 | 0.365 |
| FBAR4-7 | 0.685 | 0.847 | 0.904 | 0.677 | 0.516 | 0.633 | 0.441 | 0.530 | 0.410 | 0.375 | 0.255 |
|  | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 |
| 3 | 0.260 | 0.308 | 0.204 | 0.233 | 0.295 | 0.702 | 0.320 | 0.132 | 0.026 | 0.045 | 0.066 |
| 4 | 0.378 | 0.586 | 0.330 | 0.572 | 0.626 | 1.254 | 0.602 | 0.466 | 0.280 | 0.153 | 0.121 |
| 5 | 1.063 | 0.983 | 0.413 | 0.509 | 0.632 | 0.912 | 0.872 | 0.884 | 0.617 | 0.498 | 0.320 |
| 6 | 0.951 | 0.476 | 0.694 | 0.443 | 0.704 | 0.536 | 0.429 | 0.929 | 0.677 | 0.731 | 0.582 |
| 7 | 0.551 | 0.296 | 0.591 | 0.597 | 0.801 | 0.632 | 0.791 | 0.483 | 0.398 | 0.533 | 0.392 |
| 8 | 0.581 | 0.271 | 0.480 | 0.348 | 0.875 | 0.533 | 0.445 | 0.681 | 0.637 | 0.489 | 0.337 |
| 9 | 0.696 | 0.275 | 0.803 | 0.200 | 0.811 | 0.555 | 0.662 | 0.488 | 0.698 | 0.431 | 0.441 |
| 10 | 0.615 | 0.283 | 0.630 | 0.384 | 0.838 | 0.578 | 0.638 | 0.556 | 0.583 | 0.488 | 0.393 |
| 11+ | 0.615 | 0.283 | 0.630 | 0.384 | 0.838 | 0.578 | 0.638 | 0.556 | 0.583 | 0.488 | 0.393 |
| FBAR4-7 | 0.736 | 0.585 | 0.507 | 0.530 | 0.691 | 0.834 | 0.673 | 0.691 | 0.493 | 0.479 | 0.354 |
|  | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| 3 | 0.165 | 0.124 | 0.118 | 0.062 | 0.049 | 0.032 | 0.094 | 0.033 | 0.048 | 0.063 | 0.023 |
| 4 | 0.317 | 0.226 | 0.242 | 0.438 | 0.462 | 0.165 | 0.168 | 0.156 | 0.167 | 0.169 | 0.146 |
| 5 | 0.280 | 0.405 | 0.188 | 0.370 | 1.004 | 0.502 | 0.330 | 0.101 | 0.214 | 0.294 | 0.324 |
| 6 | 0.403 | 0.214 | 0.392 | 0.433 | 0.408 | 1.103 | 0.548 | 0.128 | 0.222 | 0.384 | 0.541 |
| 7 | 0.222 | 0.276 | 0.542 | 0.734 | 0.703 | 0.294 | 0.486 | 0.247 | 0.213 | 0.310 | 0.476 |
| 8 | 0.514 | 0.381 | 0.450 | 0.456 | 0.809 | 0.464 | 0.168 | 0.224 | 0.245 | 0.233 | 0.342 |
| 9 | 0.477 | 0.175 | 0.480 | 0.798 | 0.480 | 0.356 | 0.004 | 0.228 | 0.184 | 0.357 | 0.226 |
| 10 | 0.407 | 0.279 | 0.494 | 0.669 | 0.670 | 0.374 | 0.220 | 0.234 | 0.215 | 0.302 | 0.349 |
| 11+ | 0.407 | 0.279 | 0.494 | 0.669 | 0.670 | 0.374 | 0.220 | 0.234 | 0.215 | 0.302 | 0.349 |
| FBAR4-7 | 0.305 | 0.280 | 0.341 | 0.494 | 0.644 | 0.516 | 0.383 | 0.158 | 0.204 | 0.290 | 0.372 |

Table 4.13 (continued).

|  | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.013 | 0.024 | 0.024 | 0.025 | 0.033 | 0.084 | 0.0203 | 0.0398 |
| 4 | 0.109 | 0.091 | 0.123 | 0.143 | 0.196 | 0.220 | 0.2133 | 0.0913 |
| 5 | 0.453 | 0.258 | 0.344 | 0.376 | 0.384 | 0.460 | 0.2789 | 0.3735 |
| 6 | 0.630 | 0.466 | 0.491 | 0.614 | 0.489 | 0.476 | 0.274 | 0.3864 |
| 7 | 0.580 | 0.732 | 0.714 | 0.747 | 0.575 | 0.508 | 0.3491 | 0.2489 |
| 8 | 0.486 | 0.428 | 0.927 | 0.764 | 0.763 | 0.575 | 0.3567 | 0.2759 |
| 9 | 0.483 | 0.364 | 0.694 | 0.679 | 0.744 | 0.589 | 0.2671 | 0.457 |
| 10 | 0.524 | 0.514 | 0.782 | 0.742 | 0.700 | 0.533 | 0.3224 | 0.3263 |
| 11+ | 0.524 | 0.514 | 0.782 | 0.742 | 0.700 | 0.533 | 0.3224 | 0.3263 |
| FBAR4-7 | 0.443 | 0.387 | 0.418 | 0.470 | 0.411 | 0.416 | 0.2788 | 0.275 |
|  | 2005 | 2006 | 2007 | 2008 | 2009 | FBAR2007-2009 |  |  |
| 3 | 0.039 | 0.0277 | 0.0585 | 0.0174 | 0.024 | 0.0333 |  |  |
| 4 | 0.165 | 0.1845 | 0.1249 | 0.1181 | 0.0975 | $0.1135$ |  |  |
| 5 | 0.351 | 0.4019 | 0.404 | 0.2051 | 0.2155 | 0.2749 |  |  |
| 6 | 0.6317 | 0.4292 | 0.53 | 0.5595 | 0.305 | 0.4648 |  |  |
| 7 | 0.8285 | 0.6346 | 0.6289 | 0.5511 | 0.6244 | 0.6014 |  |  |
| 8 | 0.4734 | 0.8024 | 0.7613 | 0.6055 | 0.467 | 0.6113 |  |  |
| 9 | 0.716 | 0.4927 | 0.5996 | 0.6987 | 0.4705 | 0.5896 |  |  |
| 10 | 0.6765 | 0.8062 | 0.5427 | 0.6182 | 0.4859 | 0.5489 |  |  |
| 11+ | 0.6765 | 0.8062 | 0.5427 | 0.6182 | 0.4859 |  |  |  |
| FBAR4-7 | 0.494 | 0.4125 | 0.422 | 0.3584 | 0.3106 |  |  |  |

Table 4.14. Northeast Arctic haddock. Stock numbers at age (start of year). Numbers ' 000

| age | 1950 | 1951 | 1952 | 1953 | 1954 | 1955 | 1956 | 1957 | 1958 | 1959 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 78933 | 651953 | 71406 | 1222518 | 144953 | 61072 | 199487 | 61963 | 81070 | 384274 |
| 4 | 97323 | 54162 | 414000 | 46343 | 826025 | 98865 | 43021 | 129774 | 42882 | 56943 |
| 5 | 70978 | 43365 | 34766 | 192868 | 25153 | 517100 | 68858 | 28833 | 80884 | 28730 |
| 6 | 37681 | 25144 | 18563 | 15604 | 90908 | 14847 | 255442 | 41896 | 15950 | 36556 |
| 7 | 46516 | 13635 | 8220 | 6208 | 7787 | 49017 | 7567 | 92336 | 22753 | 7717 |
| 8 | 16065 | 11963 | 5003 | 2484 | 2488 | 3450 | 14550 | 3316 | 33416 | 7094 |
| 9 | 4591 | 4830 | 3596 | 1167 | 1056 | 859 | 1517 | 4671 | 1731 | 11465 |
| 10 | 1975 | 1968 | 948 | 742 | 572 | 221 | 458 | 835 | 2041 | 674 |
| 11+ | 5287 | 2201 | 1348 | 2339 | 957 | 218 | 418 | 408 | 1126 | 1168 |
| TOTAL | 359349 | 809220 | 557850 | 1490273 | 1099899 | 745649 | 591318 | 364031 | 281852 | 534619 |

Table 4.14 (continued).

| age | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 282781 | 127726 | 281125 | 323637 | 376654 | 119919 | 281063 | 349446 | 21122 | 20770 |
| 4 | 259615 | 169862 | 78928 | 169014 | 209012 | 252389 | 81361 | 180138 | 238266 | 14661 |
| 5 | 38174 | 142674 | 84037 | 35087 | 69225 | 121809 | 158809 | 44349 | 105992 | 128669 |
| 6 | 16493 | 18317 | 57421 | 23461 | 11090 | 27919 | 61440 | 70478 | 23389 | 47841 |
| 7 | 17055 | 7004 | 7040 | 16168 | 6827 | 3772 | 11306 | 23751 | 34098 | 12026 |
| 8 | 3465 | 8321 | 2497 | 2867 | 4861 | 2398 | 1569 | 4052 | 11418 | 13803 |
| 9 | 3784 | 1408 | 2825 | 831 | 1226 | 1522 | 1082 | 758 | 1855 | 4560 |
| 10 | 4032 | 980 | 440 | 709 | 174 | 250 | 435 | 490 | 423 | 926 |
| 11+ | 1201 | 2624 | 1350 | 638 | 1040 | 1609 | 550 | 751 | 657 | 316 |
| TOTAL year | 626599 | 478916 | 515662 | 572411 | 680109 | 531587 | 597613 | 674211 | 437221 | 243572 |
| age | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| 3 | 194165 | 112759 | 1187932 | 314585 | 61848 | 56906 | 64568 | 130778 | 203482 | 166838 |
| 4 | 13671 | 119992 | 79561 | 659854 | 166659 | 36340 | 32480 | 34642 | 46717 | 106446 |
| 5 | 9870 | 8653 | 73437 | 43204 | 291144 | 94986 | 16265 | 13770 | 7836 | 20286 |
| 6 | 62961 | 6187 | 5794 | 20281 | 12921 | 153975 | 45655 | 6915 | 4421 | 2620 |
| 7 | 21755 | 30965 | 4201 | 1819 | 10247 | 5247 | 80336 | 18360 | 3289 | 2341 |
| 8 | 6571 | 10482 | 16956 | 1982 | 1107 | 4645 | 2364 | 29528 | 7991 | 1221 |
| 9 | 6834 | 3559 | 5817 | 7768 | 1237 | 561 | 2685 | 807 | 14184 | 4194 |
| 10 | 2260 | 3771 | 2166 | 2375 | 4829 | 454 | 376 | 977 | 380 | 5989 |
| 11+ | 887 | 1916 | 3930 | 2606 | 4367 | 3209 | 3078 | 943 | 926 | 828 |
| TOT AL | 318973 | 298283 | 1379795 | 1054473 | 554359 | 356322 | 247807 | 236720 | 289226 | 310764 |
| age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 3 | 28873 | 12890 | 16184 | 9099 | 12123 | 293126 | 529534 | 116578 | 55306 | 26519 |
| 4 | 105315 | 20270 | 8874 | 10911 | 5560 | 8706 | 213186 | 260254 | 90894 | 35702 |
| 5 | 52965 | 63119 | 13789 | 6234 | 6303 | 3630 | 5594 | 112625 | 134203 | 63127 |
| 6 | 6700 | 22858 | 30681 | 8009 | 3768 | 3442 | 2462 | 3164 | 33797 | 66336 |
| 7 | 842 | 2767 | 8946 | 13936 | 4352 | 2490 | 1905 | 1307 | 1722 | 9185 |
| 8 | 1182 | 463 | 1330 | 4949 | 9143 | 2703 | 1186 | 749 | 530 | 1051 |
| 9 | 506 | 512 | 232 | 778 | 2424 | 5116 | 1412 | 615 | 273 | 273 |
| 10 | 2107 | 206 | 272 | 122 | 395 | 1666 | 2593 | 521 | 312 | 156 |
| 11+ | 3613 | 2730 | 2191 | 957 | 348 | 789 | 2209 | 1649 | 1568 | 984 |
| TOT AL | 202103 | 125815 | 82501 | 54996 | 44415 | 321668 | 760080 | 497462 | 318605 | 203334 |
| age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 3 | 36505 | 105072 | 210307 | 686563 | 302645 | 99077 | 104715 | 118575 | 59100 | 230516 |
| 4 | 19760 | 25652 | 82025 | 160733 | 516523 | 222149 | 68476 | 47787 | 71956 | 45063 |
| 5 | 24719 | 13844 | 17777 | 56691 | 110890 | 372664 | 140335 | 44943 | 32458 | 46046 |
| 6 | 37143 | 18300 | 9148 | 10846 | 31367 | 57062 | 212148 | 79483 | 24675 | 17734 |
| 7 | 31407 | 26749 | 11998 | 5099 | 5167 | 13669 | 29082 | 103935 | 34874 | 12391 |
| 8 | 4627 | 20085 | 17708 | 7204 | 2594 | 2369 | 5385 | 11662 | 40325 | 16065 |
| 9 | 727 | 3028 | 12867 | 11488 | 4192 | 1307 | 1265 | 1745 | 4448 | 15399 |
| 10 | 222 | 474 | 2062 | 7370 | 7501 | 2117 | 743 | 517 | 725 | 1730 |
| 11+ | 374 | 307 | 354 | 1202 | 5657 | 4290 | 3808 | 1557 | 1062 | 1180 |
| TOTAL | 155485 | 213511 | 364247 | 947195 | 986537 | 774705 | 565957 | 410204 | 269624 | 386124 |

Table 4.14 (continued).

| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 84921 | 370461 | 342166 | 221767 | 237476 | 370774 | 185417 | 609743 | 1028578 | 810740 |
| 4 | 173199 | 66265 | 287123 | 239328 | 140384 | 148026 | 232529 | 144661 | 452043 | 676291 |
| 5 | 29620 | 113630 | 49451 | 190955 | 156323 | 87762 | 92894 | 154700 | 102444 | 309487 |
| 6 | 23796 | 18203 | 64033 | 30100 | 104181 | 88371 | 47958 | 49985 | 82625 | 58646 |
| 7 | 9024 | 14752 | 10127 | 32367 | 14024 | 47993 | 37690 | 25194 | 23855 | 36606 |
| 8 | 6104 | 5211 | 9417 | 5937 | 13343 | 7013 | 17160 | 16360 | 10998 | 11256 |
| 9 | 7403 | 3498 | 3238 | 6260 | 3379 | 6478 | 3577 | 6298 | 6256 | 4914 |
| 10 | 6994 | 4640 | 1813 | 2215 | 3155 | 2226 | 2592 | 1789 | 2831 | 2547 |
| 11+ | 1740 | 4745 | 4136 | 3113 | 2245 | 3383 | 1413 | 2580 | 1549 | 2730 |
| TOT AL | 342802 | 601406 | 771504 | 732041 | 674511 | 762026 | 621230 | 1011310 | 1711179 | 1913218 |
|  |  | GMST | AMST |  |  |  |  |  |  |  |
| age |  | 50-** | 50-** |  |  |  |  |  |  |  |
| 3 | 0 | 131922 | 226065 |  |  |  |  |  |  |  |
| 4 | 507912 | 84057 | 141556 |  |  |  |  |  |  |  |
| 5 | 469995 | 49141 | 82709 |  |  |  |  |  |  |  |
| 6 | 190963 | 23778 | 40309 |  |  |  |  |  |  |  |
| 7 | 34029 | 11034 | 18834 |  |  |  |  |  |  |  |
| 8 | 16056 | 4981 | 8100 |  |  |  |  |  |  |  |
| 9 | 5780 | 2274 | 3650 |  |  |  |  |  |  |  |
|  | 2515 | 1013 | 1726 |  |  |  |  |  |  |  |
|  | 2659 |  |  |  |  |  |  |  |  |  |
|  | 1229908 |  |  |  |  |  |  |  |  |  |

Table 4.15. Northeast Arctic haddock. Spawning stock numbers at age (spawning time). Numbers '000

| age | 1950 | 1951 | 1952 | 1953 | 1954 | 1955 | 1956 | 1957 | 1958 | 1959 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 2684 | 22166 | 2428 | 41566 | 4928 | 2076 | 6783 | 2107 | 2756 | 13065 |
| 4 | 10608 | 5904 | 45126 | 5051 | 90037 | 10776 | 4689 | 14145 | 4674 | 6207 |
| 5 | 21152 | 12923 | 10360 | 57475 | 7495 | 154096 | 20520 | 8592 | 24103 | 8561 |
| 6 | 21818 | 14558 | 10748 | 9035 | 52636 | 8597 | 147901 | 24258 | 9235 | 21166 |
| 7 | 37492 | 10989 | 6626 | 5003 | 6276 | 39508 | 6099 | 74423 | 18339 | 6220 |
| 8 | 14877 | 11078 | 4633 | 2300 | 2304 | 3195 | 13473 | 3070 | 30943 | 6569 |
| 9 | 4476 | 4709 | 3507 | 1138 | 1030 | 838 | 1479 | 4554 | 1688 | 11178 |
| 10 | 1959 | 1952 | 941 | 736 | 567 | 219 | 454 | 828 | 2024 | 668 |
| 11+ | 5287 | 2201 | 1348 | 2339 | 957 | 218 | 418 | 408 | 1126 | 1168 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 |
| 3 | 9615 | 4343 | 9558 | 11004 | 12806 | 4077 | 9556 | 11881 | 718 | 706 |
| 4 | 28298 | 18515 | 8603 | 18423 | 22782 | 27510 | 8868 | 19635 | 25971 | 1598 |
| 5 | 11376 | 42517 | 25043 | 10456 | 20629 | 36299 | 47325 | 13216 | 31586 | 38343 |
| 6 | 9549 | 10606 | 33247 | 13584 | 6421 | 16165 | 35574 | 40807 | 13543 | 27700 |
| 7 | 13746 | 5645 | 5674 | 13031 | 5502 | 3040 | 9112 | 19143 | 27483 | 9693 |
| 8 | 3208 | 7705 | 2312 | 2655 | 4501 | 2221 | 1453 | 3752 | 10573 | 12781 |
| 9 | 3689 | 1373 | 2754 | 810 | 1196 | 1484 | 1055 | 739 | 1809 | 4446 |
| 10 | 4000 | 973 | 436 | 703 | 173 | 248 | 431 | 486 | 420 | 919 |
| 11+ | 1201 | 2624 | 1350 | 638 | 1040 | 1609 | 550 | 751 | 657 | 316 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| 3 | 6602 | 3834 | 40390 | 10696 | 2103 | 1935 | 2195 | 4446 | 6918 | 5672 |
| 4 | 1490 | 13079 | 8672 | 71924 | 18166 | 3961 | 3540 | 3776 | 5092 | 11603 |
| 5 | 2941 | 2579 | 21884 | 12875 | 86761 | 28306 | 4847 | 4103 | 2335 | 6045 |
| 6 | 36454 | 3582 | 3355 | 11743 | 7481 | 89152 | 26434 | 4004 | 2560 | 1517 |
| 7 | 17534 | 24958 | 3386 | 1466 | 8259 | 4229 | 64751 | 14798 | 2651 | 1887 |
| 8 | 6084 | 9706 | 15701 | 1835 | 1025 | 4302 | 2189 | 27343 | 7400 | 1131 |
| 9 | 6663 | 3470 | 5671 | 7573 | 1206 | 547 | 2618 | 787 | 13829 | 4089 |
| 10 | 2242 | 3741 | 2149 | 2356 | 4790 | 450 | 373 | 969 | 376 | 5941 |
| 11+ | 887 | 1916 | 3930 | 2606 | 4367 | 3209 | 3078 | 943 | 926 | 828 |

Table 4.14 (continued).

| age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 895 | 941 | 1149 | 664 | 679 | 8208 | 11650 | 2798 | 1604 | 1034 |
| 4 | 7161 | 2108 | 1766 | 2269 | 1201 | 1454 | 20039 | 18999 | 6999 | 3356 |
| 5 | 10010 | 14707 | 4109 | 3011 | 3139 | 1837 | 2428 | 29282 | 28854 | 14519 |
| 6 | 3879 | 10035 | 14666 | 4782 | 2928 | 2657 | 1856 | 2218 | 17777 | 31708 |
| 7 | 681 | 2236 | 6155 | 10146 | 3568 | 2281 | 1743 | 1191 | 1526 | 7174 |
| 8 | 1098 | 429 | 1233 | 4291 | 8119 | 2525 | 1152 | 727 | 515 | 1010 |
| 9 | 494 | 500 | 227 | 760 | 2308 | 4916 | 1381 | 610 | 270 | 270 |
| 10 | 2092 | 205 | 270 | 121 | 392 | 1640 | 2559 | 516 | 311 | 156 |
| 11+ | 3613 | 2730 | 2191 | 957 | 348 | 789 | 2209 | 1649 | 1568 | 984 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 3 | 2117 | 5149 | 7361 | 12358 | 4842 | 1585 | 2304 | 3557 | 2364 | 12678 |
| 4 | 2549 | 4592 | 12550 | 18002 | 32541 | 11774 | 4040 | 3536 | 7052 | 5723 |
| 5 | 6921 | 4693 | 7804 | 21712 | 33822 | 70061 | 24418 | 7640 | 7401 | 13169 |
| 6 | 19166 | 10431 | 5937 | 7820 | 21173 | 33781 | 90375 | 32747 | 10290 | 9027 |
| 7 | 23461 | 20623 | 9754 | 4334 | 4609 | 11864 | 23964 | 72339 | 23610 | 8475 |
| 8 | 4247 | 18096 | 16167 | 6715 | 2457 | 2284 | 5142 | 10904 | 35284 | 13848 |
| 9 | 718 | 2946 | 12430 | 11155 | 4096 | 1285 | 1251 | 1720 | 4355 | 14722 |
| 10 | 222 | 472 | 2043 | 7289 | 7426 | 2103 | 740 | 515 | 722 | 1718 |
| 11+ | 374 | 307 | 354 | 1202 | 5657 | 4290 | 3808 | 1557 | 1062 | 1180 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 3 | 2972 | 12966 | 8212 | 5544 | 6887 | 12236 | 5192 | 7927 | 9257 | 4054 |
| 4 | 28578 | 7223 | 32445 | 20343 | 11231 | 13618 | 24183 | 12730 | 22602 | 22994 |
| 5 | 10071 | 45793 | 14934 | 58623 | 36423 | 20624 | 24245 | 45791 | 26738 | 49827 |
| 6 | 13683 | 11595 | 44631 | 18030 | 61883 | 44981 | 24842 | 27442 | 48336 | 30965 |
| 7 | 6786 | 12052 | 8658 | 28613 | 11514 | 39546 | 28833 | 19550 | 19013 | 30017 |
| 8 | 5274 | 4716 | 8786 | 5628 | 12823 | 6585 | 16113 | 14920 | 10052 | 10423 |
| 9 | 7040 | 3334 | 3138 | 6122 | 3322 | 6394 | 3505 | 6166 | 6069 | 4777 |
| 10 | 6896 | 4566 | 1784 | 2190 | 3132 | 2215 | 2582 | 1778 | 2814 | 2522 |
| 11+ | 1740 | 4745 | 4136 | 3113 | 2245 | 3383 | 1413 | 2580 | 1549 | 2730 |

Table 4.16. Northeast Arctic haddock. Stock biomass at age with SOP (start of year). Tonnes

| Age \year | 1950 | 1951 | 1952 | 1953 | 1954 | 1955 | 1956 | 1957 | 1958 | 1959 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 16936 | 181121 | 13962 | 291811 | 33461 | 13543 | 53704 | 16934 | 24606 | 139214 |
| 4 | 38772 | 27938 | 150301 | 20539 | 354037 | 40708 | 21504 | 65853 | 24166 | 38303 |
| 5 | 44292 | 35037 | 19770 | 133890 | 16886 | 333509 | 53912 | 22917 | 71398 | 30270 |
| 6 | 33151 | 28642 | 14882 | 15272 | 86044 | 13501 | 281964 | 46949 | 19850 | 54301 |
| 7 | 53850 | 20437 | 8672 | 7994 | 9698 | 58650 | 10990 | 136154 | 37260 | 15085 |
| 8 | 23280 | 22446 | 6606 | 4004 | 3879 | 5167 | 26453 | 6120 | 68496 | 17356 |
| 9 | 7993 | 10888 | 5706 | 2261 | 1979 | 1546 | 3314 | 10358 | 4264 | 33704 |
| 10 | 4005 | 5168 | 1753 | 1675 | 1248 | 463 | 1166 | 2157 | 5855 | 2307 |
| 11+ | 12193 | 6572 | 2832 | 6002 | 2374 | 519 | 1211 | 1198 | 3673 | 4546 |
| TOTBIO | 234472 | 338248 | 224485 | 483448 | 509606 | 467606 | 454218 | 308640 | 259568 | 335084 |
| Age \year | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 |
| 3 | 92457 | 43715 | 90954 | 96160 | 94527 | 35506 | 82305 | 119046 | 7210 | 8022 |
| 4 | 157604 | 107943 | 47413 | 93242 | 97395 | 138748 | 44237 | 113944 | 151017 | 10513 |
| 5 | 36299 | 142014 | 79074 | 30319 | 50526 | 104888 | 135250 | 43940 | 105227 | 144527 |
| 6 | 22110 | 25705 | 76174 | 28582 | 11412 | 33894 | 73771 | 98447 | 32738 | 75762 |
| 7 | 30086 | 12933 | 12288 | 25919 | 9244 | 6025 | 17863 | 43655 | 62802 | 25060 |
| 8 | 7650 | 19233 | 5455 | 5753 | 8239 | 4795 | 3103 | 9322 | 26322 | 36002 |
| 9 | 10038 | 3910 | 7417 | 2003 | 2497 | 3657 | 2571 | 2095 | 5139 | 14290 |
| 10 | 12461 | 3172 | 1344 | 1990 | 413 | 701 | 1203 | 1576 | 1366 | 3381 |
| 11+ | 4220 | 9654 | 4695 | 2037 | 2806 | 5119 | 1730 | 2748 | 2410 | 1314 |
| TOTBIO | 372926 | 368279 | 324814 | 286005 | 277061 | 333334 | 362033 | 434775 | 394231 | 318870 |
| Age \year | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| 3 | 67683 | 50257 | 371924 | 91847 | 23557 | 21605 | 19639 | 40936 | 75286 | 73960 |
| 4 | 8848 | 99300 | 46250 | 357704 | 117861 | 25617 | 18343 | 20134 | 32093 | 87616 |
| 5 | 10006 | 11216 | 66868 | 36685 | 322509 | 104881 | 14388 | 12535 | 8432 | 26154 |
| 6 | 89990 | 11306 | 7438 | 24279 | 20179 | 239698 | 56939 | 8875 | 6707 | 4763 |
| 7 | 40915 | 74464 | 7097 | 2865 | 21058 | 10748 | 131839 | 31008 | 6565 | 5599 |
| 8 | 15468 | 31552 | 35853 | 3908 | 2848 | 11911 | 4856 | 62423 | 19969 | 3657 |
| 9 | 19330 | 12872 | 14778 | 18403 | 3824 | 1728 | 6627 | 2050 | 42584 | 15086 |
| 10 | 7447 | 15886 | 6412 | 6556 | 17385 | 1628 | 1081 | 2890 | 1327 | 25097 |
| 11+ | 3322 | 9177 | 13224 | 8177 | 17876 | 13093 | 10061 | 3174 | 3681 | 3947 |
| TOTBIO | 263010 | 316029 | 569844 | 550424 | 547098 | 430910 | 263774 | 184025 | 196644 | 245878 |
| Age \year | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 3 | 16960 | 10978 | 13841 | 4737 | 4246 | 107482 | 152058 | 37008 | 19500 | 11123 |
| 4 | 114250 | 22550 | 13201 | 11652 | 5135 | 5992 | 132975 | 146878 | 53321 | 23521 |
| 5 | 73797 | 110171 | 23307 | 9814 | 10121 | 5503 | 5532 | 116113 | 119304 | 59476 |
| 6 | 13392 | 46852 | 73332 | 13174 | 8007 | 8147 | 4761 | 4637 | 48920 | 85571 |
| 7 | 2267 | 7552 | 23667 | 30179 | 9165 | 7226 | 5334 | 3434 | 3276 | 17759 |
| 8 | 4193 | 1607 | 4438 | 11407 | 24063 | 7512 | 3847 | 2696 | 1690 | 2520 |
| 9 | 2013 | 2220 | 945 | 2169 | 6609 | 17086 | 4286 | 2455 | 1144 | 1040 |
| 10 | 9377 | 987 | 1329 | 402 | 1263 | 5660 | 9184 | 1915 | 1399 | 759 |
| 11+ | 17832 | 14376 | 11643 | 3659 | 1275 | 3061 | 7858 | 6916 | 6421 | 4964 |
| TOTBIO | 254081 | 217294 | 165702 | 87193 | 69883 | 167669 | 325835 | 322051 | 254976 | 206733 |
| Age \year | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 3 | 13726 | 38993 | 71913 | 180687 | 74340 | 26685 | 29754 | 34828 | 19284 | 74035 |
| 4 | 14413 | 17135 | 58608 | 98382 | 252723 | 99938 | 34672 | 24438 | 40625 | 26872 |
| 5 | 24429 | 15226 | 19452 | 61317 | 105755 | 283296 | 101274 | 34988 | 27312 | 40443 |
| 6 | 47894 | 24925 | 14829 | 16175 | 47351 | 75230 | 231941 | 79556 | 28594 | 21063 |
| 7 | 51719 | 44760 | 22705 | 10472 | 10005 | 26384 | 50667 | 146018 | 48447 | 18964 |
| 8 | 10725 | 40753 | 39440 | 16605 | 6553 | 5565 | 12965 | 24454 | 74281 | 28000 |
| 9 | 1999 | 8240 | 33452 | 30210 | 11556 | 3847 | 3563 | 4839 | 11616 | 34032 |
| 10 | 930 | 1475 | 6910 | 21959 | 23064 | 6666 | 2541 | 1627 | 2418 | 5198 |
| 11+ | 1937 | 1399 | 1317 | 4469 | 19250 | 14811 | 13729 | 5796 | 3922 | 4404 |
| TOTBIO | 167771 | 192905 | 268626 | 440276 | 550597 | 542421 | 481106 | 356545 | 256498 | 253011 |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 3 | 23329 | 106987 | 90696 | 59200 | 68025 | 114555 | 60079 | 202376 | 301487 | 236044 |
| 4 | 101907 | 35129 | 150509 | 117205 | 69822 | 79962 | 134076 | 86541 | 275925 | 365380 |
| 5 | 27556 | 108545 | 40595 | 156484 | 121149 | 70237 | 80017 | 140771 | 96067 | 294469 |
| 6 | 29509 | 24851 | 84612 | 34946 | 121415 | 99349 | 55305 | 61470 | 106444 | 77275 |
| 7 | 14144 | 25040 | 17848 | 55813 | 21600 | 75193 | 56815 | 38709 | 38806 | 61859 |
| 8 | 11749 | 10661 | 19633 | 12998 | 28705 | 13812 | 34104 | 31177 | 21280 | 22873 |
| 9 | 15620 | 8481 | 7856 | 15617 | 8901 | 16999 | 8609 | 15153 | 14424 | 11441 |
| 10 | 18069 | 11976 | 5072 | 6258 | 9200 | 6954 | 7999 | 5065 | 7964 | 6852 |
| 11+ | 5898 | 14567 | 12036 | 9928 | 7261 | 11445 | 5079 | 9072 | 5013 | 8716 |
| TOTBIO | 247780 | 346238 | 428856 | 468449 | 456078 | 488506 | 442085 | 590335 | 867410 | 1084910 |

Table 4.17. Northeast Arctic haddock. Spawning stock biomass at age with SOP (spawning time). Tonnes

| Age \year | 1950 | 1951 | 1952 | 1953 | 1954 | 1955 | 1956 | 1957 | 1958 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 3 | 576 | 6158 | 475 | 9922 | 1138 | 460 | 1826 | 576 | 837 |
| 4 | 4226 | 3045 | 16383 | 2239 | 38590 | 4437 | 2344 | 7178 | 2634 |
| 5 | 13199 | 10441 | 5891 | 39899 | 5032 | 99386 | 16066 | 6829 | 21276 |
| 6 | 19195 | 16584 | 8617 | 8843 | 49820 | 7817 | 163257 | 27183 | 11493 |
| 7 | 43403 | 16472 | 6990 | 6444 | 7817 | 47272 | 8858 | 109740 | 30032 |
| 8 | 21557 | 20785 | 6118 | 3708 | 3592 | 4785 | 24495 | 5667 | 63428 |
| 9 | 7793 | 10615 | 5564 | 2204 | 1929 | 1507 | 3231 | 10099 | 4157 |
| 10 | 3973 | 5126 | 1739 | 1661 | 1238 | 459 | 1156 | 2140 | 5808 |
| $11+$ | 12193 | 6572 | 2832 | 6002 | 2374 | 519 | 1211 | 1198 | 3673 |
| TO |  |  |  |  |  |  |  |  |  |

Table 4.18. Northeast Arctic haddock. Summary.

| YEAR | RECR_a3 | TOTBIO | TOTSPB | LANDINGS | YIELDSSB | SOPCOFAC | FBAR4_7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 78933 | 234472 | 126115 | 132125 | 1.0477 | 0.6148 | 0.8393 |
| 1951 | 651953 | 338248 | 95799 | 120077 | 1.2534 | 0.796 | 0.6373 |
| 1952 | 71406 | 224485 | 54608 | 127660 | 2.3378 | 0.5603 | 0.7485 |
| 1953 | 1222518 | 483448 | 80920 | 123920 | 1.5314 | 0.6839 | 0.5273 |
| 1954 | 144953 | 509606 | 111529 | 156788 | 1.4058 | 0.6614 | 0.391 |
| 1955 | 61072 | 467606 | 166642 | 202286 | 1.2139 | 0.6354 | 0.5231 |
| 1956 | 199487 | 454218 | 222445 | 213924 | 0.9617 | 0.7714 | 0.4691 |
| 1957 | 61963 | 308640 | 170611 | 123583 | 0.7244 | 0.7831 | 0.4571 |
| 1958 | 81070 | 259568 | 143338 | 112672 | 0.7861 | 0.8697 | 0.5557 |
| 1959 | 384274 | 335084 | 117294 | 88211 | 0.752 | 1.038 | 0.4137 |
| 1960 | 282781 | 372926 | 101643 | 154651 | 1.5215 | 0.9368 | 0.511 |
| 1961 | 127726 | 368279 | 115302 | 193224 | 1.6758 | 0.9807 | 0.6846 |
| 1962 | 281125 | 324814 | 104145 | 187408 | 1.7995 | 0.927 | 0.8473 |
| 1963 | 323637 | 286005 | 71199 | 146224 | 2.0537 | 0.8514 | 0.9043 |
| 1964 | 376654 | 277061 | 56226 | 99158 | 1.7636 | 0.7191 | 0.6774 |
| 1965 | 119919 | 333334 | 85889 | 118578 | 1.3806 | 0.8484 | 0.5164 |
| 1966 | 281063 | 362033 | 113339 | 161778 | 1.4274 | 0.8391 | 0.6331 |
| 1967 | 349446 | 434775 | 136736 | 136397 | 0.9975 | 0.9761 | 0.4414 |
| 1968 | 21122 | 394231 | 150787 | 181726 | 1.2052 | 0.9781 | 0.5295 |
| 1969 | 20770 | 318870 | 160490 | 130820 | 0.8151 | 1.1066 | 0.4099 |
| 1970 | 194165 | 263010 | 135209 | 88257 | 0.6527 | 0.9988 | 0.3753 |
| 1971 | 112759 | 316029 | 149142 | 78905 | 0.5291 | 1.2771 | 0.2545 |
| 1972 | 1187932 | 569844 | 114833 | 266153 | 2.3177 | 0.8971 | 0.736 |
| 1973 | 314585 | 550424 | 105653 | 322226 | 3.0498 | 0.8366 | 0.5852 |
| 1974 | 61848 | 547098 | 179900 | 221157 | 1.2293 | 1.0914 | 0.5071 |
| 1975 | 56906 | 430910 | 209653 | 175758 | 0.8383 | 1.0879 | 0.5303 |
| 1976 | 64568 | 263774 | 168276 | 137264 | 0.8157 | 0.8715 | 0.6905 |
| 1977 | 130778 | 184025 | 103296 | 110158 | 1.0664 | 0.8969 | 0.8335 |
| 1978 | 203482 | 196644 | 82753 | 95422 | 1.1531 | 1.0601 | 0.6732 |
| 1979 | 166838 | 245878 | 74067 | 103623 | 1.399 | 1.2702 | 0.6905 |
| 1980 | 28873 | 254081 | 64835 | 87889 | 1.3556 | 1.2854 | 0.4928 |
| 1981 | 12890 | 217294 | 74504 | 77153 | 1.0356 | 1.3583 | 0.4785 |
| 1982 | 16184 | 165702 | 79890 | 46955 | 0.5877 | 1.3511 | 0.3536 |
| 1983 | 9099 | 87193 | 53412 | 24600 | 0.4606 | 0.9535 | 0.3051 |
| 1984 | 12123 | 69883 | 50312 | 20945 | 0.4163 | 0.9491 | 0.2804 |
| 1985 | 293126 | 167669 | 51769 | 45052 | 0.8702 | 1.0242 | 0.3411 |
| 1986 | 529534 | 325835 | 51570 | 100563 | 1.95 | 0.9508 | 0.4937 |
| 1987 | 116578 | 322051 | 62044 | 154916 | 2.4969 | 1.0078 | 0.6443 |
| 1988 | 55306 | 254976 | 69547 | 95255 | 1.3696 | 1.0045 | 0.5159 |
| 1989 | 26519 | 206733 | 80270 | 58518 | 0.729 | 1.023 | 0.3829 |
| 1990 | 36505 | 167771 | 87527 | 27182 | 0.3106 | 0.9843 | 0.158 |
| 1991 | 105072 | 192905 | 106460 | 36216 | 0.3402 | 0.9639 | 0.2039 |
| 1992 | 210307 | 268626 | 124594 | 59922 | 0.4809 | 1.0207 | 0.2895 |
| 1993 | 686563 | 440276 | 129315 | 82379 | 0.637 | 0.9969 | 0.3716 |
| 1994 | 302645 | 550597 | 149832 | 135186 | 0.9023 | 0.9945 | 0.4429 |
| 1995 | 99077 | 542421 | 156997 | 142448 | 0.9073 | 0.9759 | 0.3865 |
| 1996 | 104715 | 481106 | 193041 | 178128 | 0.9227 | 0.9832 | 0.4177 |
| 1997 | 118575 | 356545 | 178260 | 154359 | 0.8659 | 0.9505 | 0.47 |
| 1998 | 59100 | 256498 | 138400 | 100630 | 0.7271 | 0.9888 | 0.4109 |
| 1999 | 230516 | 253011 | 108980 | 83195 | 0.7634 | 0.9792 | 0.4158 |
| 2000 | 84921 | 247780 | 103324 | 68944 | 0.6673 | 0.9741 | 0.2788 |
| 2001 | 370461 | 346238 | 131687 | 89640 | 0.6807 | 1.0098 | 0.275 |
| 2002 | 342166 | 428856 | 148634 | 114798 | 0.7724 | 0.989 | 0.3241 |
| 2003 | 221767 | 468449 | 173468 | 138926 | 0.8009 | 0.9814 | 0.4525 |
| 2004 | 237476 | 456078 | 178372 | 158279 | 0.8874 | 0.981 | 0.4019 |
| 2005 | 370774 | 488506 | 188282 | 158298 | 0.8408 | 0.9967 | 0.494 |
| 2006 | 185417 | 442085 | 162129 | 153157 | 0.9447 | 1.0063 | 0.4125 |
| 2007 | 609743 | 590335 | 173075 | 161525 | 0.9333 | 1.0088 | 0.422 |
| 2008 | 1028578 | 867410 | 181152 | 155604 | 0.859 | 1.0073 | 0.3584 |
| 2009 | 810740 | 1084910 | 200339 | 200512 | 1.0009 | 1.0005 | 0.3106 |
| Mean | 249185 | 360953 | 122664 | 127022 | 1.1037 | 0.9594 | 0.4863 |

Table 4.19. Northeast Arctic haddock. Prediction with management option table: Input data

| 2010 |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| 3 | 212000 | 0.362 | 0.010 | 0 | 0 | 0.302 | 0.0284 | 0.679 |
| 4 | 508063 | 0.249 | 0.040 | 0 | 0 | 0.559 | 0.0969 | 0.917 |
| 5 | 470158 | 0.281 | 0.160 | 0 | 0 | 0.866 | 0.2347 | 1.182 |
| 6 | 191039 | 0.235 | 0.432 | 0 | 0 | 1.298 | 0.3968 | 1.466 |
| 7 | 34048 | 0.2 | 0.809 | 0 | 0 | 1.661 | 0.5131 | 1.777 |
| 8 | 16076 | 0.2 | 0.932 | 0 | 0 | 2.04 | 0.5213 | 2.075 |
| 9 | 5793 | 0.2 | 0.976 | 0 | 0 | 2.438 | 0.5026 | 2.305 |
| 10 | 2523 | 0.2 | 0.992 | 0 | 0 | 2.786 | 0.4675 | 2.46 |
| 11 | 1288 | 0.2 | 1.000 | 0 | 0 | 3.205 | 0.4675 | 2.86 |

2011

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 101000 | 0.362 | 0.002 | 0 | 0 | 0.301 | 0.0284 | 0.679 |
| 4 | $\cdot$ | 0.249 | 0.074 | 0 | 0 | 0.558 | 0.0969 | 0.917 |
| 5 | $\cdot$ | 0.281 | 0.228 | 0 | 0 | 0.881 | 0.2347 | 1.182 |
| 6 | $\cdot$ | 0.235 | 0.503 | 0 | 0 | 1.245 | 0.3968 | 1.466 |
| 7 | $\cdot$ | 0.2 | 0.756 | 0 | 0 | 1.640 | 0.5131 | 1.777 |
| 8 | $\cdot$ | 0.2 | 0.901 | 0 | 0 | 2.055 | 0.5213 | 2.075 |
| 9 | $\cdot$ | 0.2 | 0.965 | 0 | 0 | 2.487 | 0.5026 | 2.305 |
| 10 | $\cdot$ | 0.2 | 0.988 | 0 | 0 | 2.938 | 0.4675 | 2.46 |
| 11 | $\cdot$ | 0.2 | 1 | 0 | 0 | 3.361 | 0.4675 | 2.86 |

2012

| Age | N | M | Mat | PF | PM | SWt | Sel | CWt |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 303000 | 0.362 | 0.002 | 0 | 0 | 0.301 | 0.0284 | 0.679 |
| 4 | . | 0.249 | 0.074 | 0 | 0 | 0.558 | 0.0969 | 0.917 |
| 5 | $\cdot$ | 0.281 | 0.228 | 0 | 0 | 0.881 | 0.2347 | 1.182 |
| 6 | $\cdot$ | 0.235 | 0.503 | 0 | 0 | 1.245 | 0.3968 | 1.466 |
| 7 | . | 0.2 | 0.756 | 0 | 0 | 1.640 | 0.5131 | 1.777 |
| 8 | . | 0.2 | 0.901 | 0 | 0 | 2.055 | 0.5213 | 2.075 |
| 9 | $\cdot$ | 0.2 | 0.965 | 0 | 0 | 2.487 | 0.5026 | 2.305 |
| 10 | $\cdot$ | 0.2 | 0.988 | 0 | 0 | 2.938 | 0.4675 | 2.46 |
| 11 | . | 0.2 | 1 | 0 | 0 | 3.361 | 0.4675 | 2.86 |

Table 4.20. Northeast Arctic haddock. Prediction with management option table for 2010-2012

| Biomass 2010 |  | SSB2009 | FMult | FBar2009 |  | Landings2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1117786 |  | 285470 | 1 | 0.3104 |  | 268616 |
| 2011 |  |  |  |  | 2012 |  |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 1011972 | 444470 | 0 | 0 | 0 | 1167816 | 689292 |
| . | 444470 | 0.1 | 0.031 | 35272 | 1132956 | 663537 |
| . | 444470 | 0.2 | 0.0621 | 69275 | 1099426 | 638834 |
| - | 444470 | 0.3 | 0.0931 | 102058 | 1067171 | 615137 |
| - | 444470 | 0.4 | 0.1242 | 133671 | 1036138 | 592404 |
| . | 444470 | 0.5 | 0.1552 | 164160 | 1006278 | 570593 |
| . | 444470 | 0.6 | 0.1862 | 193571 | 977541 | 549665 |
| - | 444470 | 0.7 | 0.2173 | 221946 | 949882 | 529581 |
| - | 444470 | 0.8 | 0.2483 | 249326 | 923257 | 510306 |
| . | 444470 | 0.9 | 0.2793 | 275751 | 897624 | 491804 |
| . | 444470 | 1 | 0.3104 | 301257 | 872941 | 474044 |
| . | 444470 | 1.1 | 0.3414 | 325882 | 849171 | 456993 |
| . | 444470 | 1.2 | 0.3725 | 349659 | 826277 | 440622 |
| . | 444470 | 1.3 | 0.4035 | 372623 | 804223 | 424902 |
| - | 444470 | 1.4 | 0.4345 | 394803 | 782974 | 409804 |
| . | 444470 | 1.5 | 0.4656 | 416232 | 762500 | 395304 |
| . | 444470 | 1.6 | 0.4966 | 436937 | 742768 | 381375 |
| . | 444470 | 1.7 | 0.5276 | 456948 | 723749 | 367993 |
| . | 444470 | 1.8 | 0.5587 | 476290 | 705415 | 355137 |
| - | 444470 | 1.9 | 0.5897 | 494989 | 687738 | 342783 |
| . | 444470 | 2 | 0.6208 | 513070 | 670692 | 330910 |

Table 4.21. Northeast Arctic haddock. Prediction single option table for 2009-2011

| Year | 2010 | F multiplier | 1 | Fbar: | 0.3104 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos | SSB |
| 3 | 0.0284 | 4985 | 3385 | 212000 | 64024 | 2120 | 640 |
| 4 | 0.0969 | 41619 | 38165 | 508063 | 284007 | 20323 | 11360 |
| 5 | 0.2347 | 86214 | 101905 | 470158 | 407157 | 75225 | 65145 |
| 6 | 0.3968 | 56195 | 82382 | 191039 | 247969 | 82529 | 107122 |
| 7 | 0.5131 | 12491 | 22197 | 34048 | 56554 | 27545 | 45752 |
| 8 | 0.5213 | 5971 | 12389 | 16076 | 32795 | 14983 | 30565 |
| 9 | 0.5026 | 2091 | 4821 | 5793 | 14123 | 5654 | 13784 |
| 10 | 0.4675 | 861 | 2117 | 2523 | 7029 | 2503 | 6973 |
| 11 | 0.4675 | 439 | 1256 | 1288 | 4128 | 1288 | 4128 |
| Total |  | 210866 | 268616 | 1440988 | 1117786 | 232169 | 285470 |
| Year | 2011 | F multiplier. | 1.1277 | Fbar: | 0.35 |  |  |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos | SSB |
| 3 | 0.032 | 2673 | 1815 | 101000 | 30401 | 202 | 61 |
| 4 | 0.1093 | 13177 | 12083 | 143479 | 80061 | 10617 | 5925 |
| 5 | 0.2647 | 73328 | 86674 | 359497 | 316717 | 81965 | 72211 |
| 6 | 0.4475 | 91040 | 133464 | 280723 | 349500 | 141204 | 175799 |
| 7 | 0.5786 | 40828 | 72551 | 101563 | 166563 | 76781 | 125922 |
| 8 | 0.5878 | 6788 | 14086 | 16688 | 34293 | 15036 | 30898 |
| 9 | 0.5668 | 3093 | 7130 | 7815 | 19436 | 7541 | 18755 |
| 10 | 0.5272 | 1075 | 2644 | 2869 | 8430 | 2835 | 8329 |
| 11 | 0.5272 | 732 | 2095 | 1955 | 6571 | 1955 | 6571 |
| Total |  | 232735 | 332542 | 1015588 | 1011972 | 338137 | 444470 |
| Year | 2012 | F multiplier. | 1.1277 | Fbar: | 0.35 |  |  |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos | SSB |
| 3 | 0.032 | 8020 | 5446 | 303000 | 91203 | 606 | 182 |
| 4 | 0.1093 | 6255 | 5736 | 68108 | 38004 | 5040 | 2812 |
| 5 | 0.2647 | 20454 | 24176 | 100275 | 88342 | 22863 | 20142 |
| 6 | 0.4475 | 67557 | 99038 | 208313 | 259350 | 104782 | 130453 |
| 7 | 0.5786 | 57032 | 101345 | 141870 | 232667 | 107254 | 175896 |
| 8 | 0.5878 | 18965 | 39352 | 46622 | 95808 | 42007 | 86323 |
| 9 | 0.5668 | 3004 | 6925 | 7590 | 18876 | 7324 | 18215 |
| 10 | 0.5272 | 1360 | 3345 | 3630 | 10665 | 3587 | 10537 |
| 11 | 0.5272 | 873 | 2498 | 2331 | 7836 | 2331 | 7836 |
| Total |  | 183520 | 287861 | 881740 | 842753 | 295793 | 452398 |

Table 4.22. Northeast Arctic haddock. Prediction using catch constraint for 2011-2012

| Year | 2010 | F multiplier: | 1 | Fbar: | 0.3104 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos | SSB |
| 3 | 0.0284 | 4985 | 3385 | 212000 | 64024 | 2120 | 640 |
| 4 | 0.0969 | 41619 | 38165 | 508063 | 284007 | 20323 | 11360 |
| 5 | 0.2347 | 86214 | 101905 | 470158 | 407157 | 75225 | 65145 |
| 6 | 0.3968 | 56195 | 82382 | 191039 | 247969 | 82529 | 107122 |
| 7 | 0.5131 | 12491 | 22197 | 34048 | 56554 | 27545 | 45752 |
| 8 | 0.5213 | 5971 | 12389 | 16076 | 32795 | 14983 | 30565 |
| 9 | 0.5026 | 2091 | 4821 | 5793 | 14123 | 5654 | 13784 |
| 10 | 0.4675 | 861 | 2117 | 2523 | 7029 | 2503 | 6973 |
| 11 | 0.4675 | 439 | 1256 | 1288 | 4128 | 1288 | 4128 |
| Total |  | 210866 | 268616 | 1440988 | 1117786 | 232169 | 285470 |

Year: 2011 Fbar=0.3125: Catch constraint 243000*1.25=303000

| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos | SSB |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 0.0286 | 2391 | 1624 | 101000 | 30401 | 202 | 61 |
| 4 | 0.0976 | 11831 | 10849 | 143479 | 80061 | 10617 | 5925 |
| 5 | 0.2363 | 66331 | 78403 | 359497 | 316717 | 81965 | 72211 |
| 6 | 0.3996 | 83049 | 121749 | 280723 | 349500 | 141204 | 175799 |
| 7 | 0.5167 | 37461 | 66569 | 101563 | 166563 | 76781 | 125922 |
| 8 | 0.5249 | 6231 | 12929 | 16688 | 34293 | 15036 | 30898 |
| 9 | 0.5061 | 2837 | 6539 | 7815 | 19436 | 7541 | 18755 |
| 10 | 0.4708 | 984 | 2421 | 2869 | 8430 | 2835 | 8329 |
| 11 | 0.4708 | 671 | 1918 | 1955 | 6571 | 1955 | 6571 |
| Total |  | 211785 | 303000 | 1015588 | 1011972 | 338137 | 444470 |

Year: 2012 Fbar=0.35: F multiplier: 1.1277

| Age | F | CatchNos | Yield | StockNos | Biomass | SSNos | SSB |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 0.032 | 8020 | 5446 | 303000 | 91203 | 606 | 182 |
| 4 | 0.1093 | 6276 | 5755 | 68342 | 38135 | 5057 | 2822 |
| 5 | 0.2647 | 20694 | 24461 | 101455 | 89382 | 23132 | 20379 |
| 6 | 0.4475 | 69498 | 101884 | 214299 | 266802 | 107792 | 134201 |
| 7 | 0.5786 | 59829 | 106317 | 148830 | 244081 | 112516 | 184526 |
| 8 | 0.5878 | 20177 | 41867 | 49601 | 101929 | 44690 | 91838 |
| 9 | 0.5668 | 3199 | 7374 | 8083 | 20102 | 7800 | 19398 |
| 10 | 0.5272 | 1445 | 3555 | 3857 | 11332 | 3811 | 11196 |
| 11 | 0.5272 | 924 | 2643 | 2467 | 8291 | 2467 | 8291 |
| Total |  | 190063 | 299301 | 899933 | 871257 | 307871 | 472834 |

Table 4.23. Northeast Arctic haddock. Yield per recruit. Input data and results.

MFYPR version 2 a
TestProjection index file.
Time and date: 15:32 26.04.2009
Fbar age range: 4-7

| Age | M | Mat | PF | PM | SWt | Sel | CWt |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 3 | 0.362 | 0.0031 | 0 | 0 | 0.324 | 0.0284 | 0.752 |
| 4 | 0.249 | 0.107 | 0 | 0 | 0.617 | 0.0969 | 1.0486 |
| 5 | 0.281 | 0.304 | 0 | 0 | 0.994 | 0.2347 | 1.3372 |
| 6 | 0.235 | 0.588 | 0 | 0 | 1.412 | 0.3968 | 1.6387 |
| 7 | 0.2 | 0.811 | 0 | 0 | 1.867 | 0.5131 | 1.8935 |
| 8 | 0.2 | 0.926 | 0 | 0 | 2.331 | 0.5213 | 2.17 |
| 9 | 0.2 | 0.975 | 0 | 0 | 2.802 | 0.5026 | 2.4398 |
| 10 | 0.2 | 0.992 | 0 | 0 | 3.274 | 0.4675 | 2.7296 |
| 11 | 0.2 | 1 | 0 | 0 | 3.737 | 0.4675 | 3.2192 |

Yield per results

| FMult | Fbar | Catch |  | Stock |  | SpwnNos |  | SpwnNos |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Nos | Yield | Nos | Biomass | Jan | SSBJan | Spwn | SSBSpwn |
| 0 | 0 | 0 | 0 | 4.4371 | 7.2894 | 2.184 | 5.7888 | 2.184 | 5.7888 |
| 0.1 | 0.031 | 0.0921 | 0.1973 | 3.9883 | 5.8098 | 1.7557 | 4.3407 | 1.7557 | 4.3407 |
| 0.2 | 0.0621 | 0.1559 | 0.3167 | 3.6814 | 4.8413 | 1.4681 | 3.4013 | 1.4681 | 3.4013 |
| 0.3 | 0.0931 | 0.2029 | 0.3936 | 3.4575 | 4.1665 | 1.2626 | 2.7538 | 1.2626 | 2.7538 |
| 0.4 | 0.1242 | 0.2393 | 0.4457 | 3.2863 | 3.6744 | 1.1088 | 2.2871 | 1.1088 | 2.2871 |
| 0.5 | 0.1552 | 0.2686 | 0.4823 | 3.1505 | 3.3024 | 0.9897 | 1.9391 | 0.9897 | 1.9391 |
| 0.6 | 0.1862 | 0.2928 | 0.5089 | 3.0398 | 3.0131 | 0.8949 | 1.6722 | 0.8949 | 1.6722 |
| 0.7 | 0.2173 | 0.3132 | 0.5289 | 2.9474 | 2.7826 | 0.8177 | 1.4628 | 0.8177 | 1.4628 |
| 0.8 | 0.2483 | 0.3308 | 0.5443 | 2.8688 | 2.5952 | 0.7537 | 1.2954 | 0.7537 | 1.2954 |
| 0.9 | 0.2793 | 0.3463 | 0.5564 | 2.8008 | 2.4402 | 0.6997 | 1.1592 | 0.6997 | 1.1592 |
| 1 | 0.3104 | 0.36 | 0.5662 | 2.7413 | 2.3099 | 0.6536 | 1.0468 | 0.6536 | 1.0468 |
| 1.1 | 0.3414 | 0.3723 | 0.5742 | 2.6885 | 2.1989 | 0.6137 | 0.9529 | 0.6137 | 0.9529 |
| 1.2 | 0.3725 | 0.3835 | 0.581 | 2.6412 | 2.1032 | 0.579 | 0.8734 | 0.579 | 0.8734 |
| 1.3 | 0.4035 | 0.3937 | 0.5867 | 2.5985 | 2.0198 | 0.5483 | 0.8054 | 0.5483 | 0.8054 |
| 1.4 | 0.4345 | 0.403 | 0.5916 | 2.5596 | 1.9464 | 0.5211 | 0.7468 | 0.5211 | 0.7468 |
| 1.5 | 0.4656 | 0.4117 | 0.596 | 2.5239 | 1.8812 | 0.4967 | 0.6957 | 0.4967 | 0.6957 |
| 1.6 | 0.4966 | 0.4198 | 0.5998 | 2.491 | 1.8229 | 0.4747 | 0.651 | 0.4747 | 0.651 |
| 1.7 | 0.5276 | 0.4274 | 0.6031 | 2.4606 | 1.7704 | 0.4549 | 0.6114 | 0.4549 | 0.6114 |
| 1.8 | 0.5587 | 0.4345 | 0.6062 | 2.4322 | 1.7228 | 0.4368 | 0.5763 | 0.4368 | 0.5763 |
| 1.9 | 0.5897 | 0.4412 | 0.6089 | 2.4057 | 1.6794 | 0.4202 | 0.5448 | 0.4202 | 0.5448 |
| 2 | 0.6208 | 0.4475 | 0.6114 | 2.3808 | 1.6396 | 0.405 | 0.5165 | 0.405 | 0.5165 |
| Reference point |  | F multiplier |  | Absolute F |  |  |  |  |  |
| Fbar(4-7) |  | 1 |  | 0.3401 |  |  |  |  |  |
| FMax |  | >=1000000 |  |  |  |  |  |  |  |
| F0.1 |  | 0.5606 |  | $0.174$ |  |  |  |  |  |
| F35\%SPR |  | 0.4723 |  | 0.1466 |  |  |  |  |  |

Weights in kilograms


Figure 4.1A Landings of Northeast Arctic haddock 1950-2009


Figure 4.1B Fishing mortality of Northeast Arctic haddock 1950-2009


Figure 4.1C Recruitment of Northeast Arctic haddock 1950-2010


Figure 4.1D Spawning stock biomass of Northeast Arctic haddock 1950-2010


Figure 4.2 Stock-Recruitment relationship of Northeast Arctic haddock 1950-2009


Figure 4.3 Yield and Spawning Stock Biomass per Recruit of Northeast Arctic haddock


Figure 4.4 Spawning stock biomass - fishing mortality relationship of Northeast Arctic haddock 1950-2009
retro SSB, tonnes

retro Fishing mortality (Fbar 4-7)

Recruitment Age3, '000


Figure 4.5. Retrospective plots for assessment years 1994-2010 using standard settings in the XSA runs and keeping weight, maturity and natural mortality as estimated in 2010 for all runs.



Figure 4.6. Northeast Arctic haddock; log catchability residuals plot, fleets combined, with shrinkage 0.5


Figure 4.6 (continued).


Figure 4.7a Northeast Arctic haddock. Sensitivit y analys is of XSA to settings plusgroup=(9,10,11) for Fishing mortality, Spawning stock biomass, and Recruitment at age 3 for the time period 1950 to 2009


Figure 4.7b Northeast Arctic haddock. Sensitivity analysis of XSA to settings F shr=(0.5,1.0,1.5) for Fishing mortality, Spawning stock biomass, and Recruitment at age 3 for the time period 1950 to 2009


Figure 4.7c Northeast Arctic haddock. Sensitivity analysis of XSA to settings $q$-plateau=( $7,8,9$ ) for Fishing mortality, Spawning stock biomass, and Recruitment at age 3 for the time period 1950 to 2009


Figure 4.7 d Northeast Arctic haddock. Sensitivity analys is of XSA to settings tspower=( $1,2,3$ ) for Fishing mortality, Spawning stock biomass, and Recruitment at age 3 for the time period 1950 to 2009


Figure 4.7e Northeast Arctic haddock. Sensitivity analysis of XSA to settings surveys=all for Fishing mortality, Spawning stock biomass, and Recruitment at age 3 for the time period 1950 to 2009

Recr age 3


SSB


F4-7


Figure 4.8 Northeast Arctic haddock. Sensitivity analysis of XSA to settings with adding a new surveys, for Fishing mortality, Spawning stock biomass, and Recruitment at age 3 for the time period 1950 to 2009

Table B1 Northeast Arctic haddock. Results from the Norwegian bottom trawl surve y in the Barents Sea in January-March. Index of number of fish at age. Indices for 1983-1998 revised August 1999.

| Year | Age |  |  |  |  |  |  |  |  |  | Total | $\begin{gathered} \text { Area } \\ \text { covered } \\ (1000 \\ \left.n m^{2}\right) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |  |  |
| 1981 | 3.1 | 7.3 | 2.3 | 7.8 | 1.8 | 5.3 | 0.5 | 0.2 | 0 | 0 | 28.3 | 88.1 |
| 1982 | 3.9 | 1.5 | 1.7 | 1.8 | 1.9 | 4.8 | 2.4 | 0.2 | 0 | 0 | 18.2 | 88.1 |
| 1983 | 2919.3 | 4.8 | 3.1 | 2.4 | 0.9 | 1.9 | 2.5 | 0.7 | 0 | 0 | 2935.6 | 88.1 |
| 1984 | 3832.6 | 514.6 | 18.9 | 1.5 | 0.8 | 0.2 | 0.1 | 0.4 | 0.1 | 0 | 4369.2 | 88.1 |
| 1985 | 1901.1 | 1593.8 | 475.9 | 14.7 | 0.5 | 0.5 | 0.1 | 0.1 | 0.4 | 0.3 | 3987.4 | 88.1 |
| 1986 | 665.0 | 370.3 | 384.6 | 110.8 | 0.6 | 0.2 | 0.1 | 0.1 | 0.1 | 0.1 | 1531.9 | 88.1 |
| 1987 | 163.8 | 79.9 | 154.4 | 290.2 | 52.9 | 0.0 | 0 | 0 | 0 | 0.3 | 741.5 | 88.1 |
| 1988 | 35.4 | 15.3 | 25.3 | 68.9 | 116.4 | 13.8 | 0.1 | 0 | 0 | 0 | 275.2 | 88.1 |
| 1989 | 81.2 | 9.5 | 14.1 | 21.6 | 34.0 | 32.7 | 3.4 | 0.1 | 0 | 0 | 196.6 | 88.1 |
| 1990 | 644.1 | 54.6 | 4.5 | 3.4 | 5.0 | 9.2 | 11.8 | 1.8 | 0 | 0 | 734.4 | 88.1 |
| 1991 | 2006.0 | 300.3 | 33.4 | 5.1 | 4.2 | 2.7 | 1.7 | 4.2 | 0 | 0 | 2357.6 | 88.1 |
| 1992 | 1659.4 | 1375.5 | 150.5 | 24.4 | 2.1 | 0.6 | 0.7 | 1.6 | 2.3 | 0 | 3217.1 | 88.1 |
| 1993 | 727.9 | 599.0 | 507.7 | 105.6 | 10.5 | 0.6 | 0.4 | 0.3 | 0.4 | 1.1 | 1953.5 | 137.6 |
| 1994 | 603.2 | 228.0 | 339.5 | 436.6 | 49.7 | 3.4 | 0.2 | 0.1 | 0.2 | 0.6 | 1661.5 | 143.8 |
| 1995 | 1463.6 | 179.3 | 53.6 | 171.1 | 339.5 | 34.5 | 2.8 | 0 | 0.1 | 0 | 2244.5 | 186.6 |
| 1996 | 309.5 | 263.6 | 52.5 | 48.1 | 148.6 | 252.8 | 11.6 | 0.9 | 0 | 0.1 | 1087.7 | 165.3 |
| 19971 | 1268.0 | 67.9 | 86.1 | 28.0 | 19.4 | 46.7 | 62.2 | 3.5 | 0.1 | 0 | 1581.9 | 87.5 |
| $1998{ }^{1}$ | 212.9 | 137.9 | 22.7 | 33.2 | 13.2 | 3.4 | 8.0 | 8.1 | 0.7 | 0.1 | 440.2 | 99.2 |
| 1999 | 1244.9 | 57.6 | 59.8 | 12.2 | 10.2 | 2.8 | 1.0 | 1.7 | 1.1 | 0 | 1391.3 | 118.3 |
| 2000 | 847.2 | 452.2 | 27.2 | 35.4 | 8.4 | 4.0 | 0.8 | 0.3 | 0.7 | 0.2 | 1376.4 | 162.4 |
| 2001 | 1220.5 | 460.3 | 296.0 | 29.3 | 25.1 | 1.7 | 0.9 | 0.1 | 0.1 | 0.3 | 2034.3 | 164.1 |
| 2002 | 1680.3 | 534.7 | 314.7 | 185.3 | 17.6 | 8.2 | 0.8 | 0.3 | 0 | 0.3 | 2742.2 | 156.7 |
| 2003 | 3332.1 | 513.1 | 317.4 | 182 | 73.6 | 5.5 | 2.3 | 0.2 | 0.1 | 0.2 | 4426.5 | 146.6 |
| 2004 | 715.9 | 711.2 | 188.1 | 102.7 | 80.4 | 46.2 | 5.9 | 1.1 | 0.2 | 0.1 | 1852 | 164.6 |
| 2005 | 4630.2 | 420.4 | 346.5 | 133.3 | 66.8 | 52.2 | 12.3 | 0.6 | 0.2 | 0 | 5662.4 | 178.9 |
| 2006 | 5141.3 | 1313.1 | 77.4 | 140.5 | 48.2 | 19.6 | 15.2 | 3.1 | 0.1 | 0.3 | 6758.8 | 1691 |
| $2007{ }^{1}$ | 3874.4 | 1593.8 | 507.7 | 66 | 86 | 23.3 | 7.5 | 3.7 | 1.4 | 0.2 | 6164 | 122.2 |
| 2008 | 860.2 | 2129.4 | 1522.4 | 600.9 | 86.8 | 48.9 | 6.27 | 2.51 | 0.82 | 0.13 | 7257 | 164.4 |
| 2009 | 564.7 | 328 | 1270.4 | 773.2 | 365.4 | 38.5 | 10.6 | 1.4 | 0.1 | 0.3 | 998 | 170.9 |
| 2010 | 1619.5 | 111.2 | 102.8 | 508.6 | 479.6 | 131.2 | 7 | 1 | 0.6 | 0.6 | 2962 | 159.9 |

[^3]Survey areas extended from 1993 onwards.

Table B2 Northeast Arctic haddock. Results from the Russian trawl survey in the Barents Sea and adjacent waters in late autumn (numbers per hour trawling).

|  | Age |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | Total |
|  |  |  |  |  | Sub-area I |  |  |  |  |  |  |  |
| 1983 | 39.9 | 97.3 | 16.5 | 0.8 | 0.7 | + | - | - | - | - | 1.1 | 156.3 |
| 1984 | 9.7 | 100.2 | 110.6 | 2.8 | 0.4 | 0.2 | + | - | - | - | 0.7 | 224.6 |
| 1985 | 3.9 | 19.1 | 213.4 | 168.8 | 0.8 | 0.2 | 0.1 | - | - | - | 0.3 | 406.6 |
| 1986 | 0.2 | 2.3 | 16.6 | 58.1 | 27.6 | 0.1 | + | + | + | - | - | 105 |
| 1987 | 0.4 | 1.4 | 2.5 | 12.5 | 34.2 | 8.6 | + | + | - | + | - | 59.8 |
| 1988 | 1.9 | 0.4 | 1.1 | 2.8 | 6.2 | 11.6 | 1.1 | + | + | + | - | 25.2 |
| 1989 | 3.3 | 3 | 3.6 | 0.7 | 2.5 | 7.1 | 13.9 | 1.8 | 0.1 | + | - | 36 |
| 1990 | 71.7 | 22.2 | 18.6 | 13.2 | 7.5 | 13.2 | 13.3 | 10.3 | 0.6 | 0.1 | - | 170.7 |
| 1991 | 15.9 | 61.5 | 27.5 | 10.8 | 1.6 | 0.6 | 1 | 3.3 | 2.6 | 0.3 | - | 125.1 |
| 1992 | 19.6 | 44.2 | 180.6 | 52.1 | 8.4 | 0.7 | 1 | 1.6 | 1.3 | 0.2 | - | 309.7 |
| 1993 | 5.5 | 8.1 | 69.2 | 371.5 | 78.4 | 10.2 | 1.4 | 0.7 | 0.8 | 1.8 | - | 547.7 |
| 1994 | 13.5 | 6.7 | 8 | 65.9 | 146 | 15.9 | 1.7 | 0.1 | 0.2 | 0.7 | - | 258.8 |
| 1995 | 9.9 | 12.7 | 6.5 | 4 | 26.8 | 77.6 | 7.3 | 1 | 0.1 | 0.5 | - | 146.3 |
| 1996 | 5 | 3.1 | 5.6 | 3.4 | 7.7 | 62.3 | 56.5 | 4.8 | 0.4 | 0.6 | - | 149.3 |
| $1997{ }^{1}$ | 2.7 | 6.9 | 3.2 | 5.3 | 5.5 | 1.5 | 4.5 | 1.7 | 1.5 | - | - | 32.7 |
| 1998 | 10.5 | 2.9 | 17.2 | 6.7 | 7.8 | 0.6 | 0.9 | 2.1 | 0.7 | + | - | 49.4 |
| 1999 | 6.9 | 34.9 | 8.8 | 34 | 5.3 | 5.6 | 1.2 | 0.3 | 0.9 | 0.3 | - | 98.2 |
| 2000 | 18 | 25.4 | 37.5 | 9.3 | 13 | 3.2 | 1.1 | 0.2 | 0.1 | 0.4 | - | 108.3 |
| 2001 | 30.5 | 18.6 | 42.3 | 58.9 | 5.8 | 6.8 | 0.8 | 0.5 | 0.1 | 0.1 | - | 164.5 |
| 2002 | 39.7 | 29.2 | 29.4 | 69.2 | 74.7 | 6.7 | 3.2 | 0.6 | 0.1 | 0.2 | - | 252.7 |
| 2003 | 28.1 | 38.9 | 35.4 | 28.1 | 43 | 28 | 3.5 | 0.8 | 0.1 | 0.1 | - | 206 |
| 2004 | 47.9 | 12 | 27.9 | 18.6 | 12.8 | 16.1 | 12.4 | 0.8 | 0.3 | 0.1 | - | 148.9 |
| 2005 | 62.7 | 109.6 | 20.7 | 34.4 | 12.4 | 6.5 | 7.1 | 2.5 | 0.1 | 0.1 | - | 256.1 |
| $2006{ }^{3}$ | 48 | 168.7 | 157.9 | 15.2 | 25.5 | 7.3 | 3.1 | 2.7 | 0.8 | 0.2 | - | 429.4 |
| 2007 | 4.3 | 90.2 | 153.6 | 98.7 | 9.1 | 9 | 2.3 | 0.7 | 0.4 | 0.1 | - | 368.5 |
| 2008 | 5.9 | 14.6 | 284.4 | 283.4 | 153 | 17.2 | 11.8 | 1.5 | 0.3 | 0.3 | - | 772.5 |
| 2009 | 14.7 | 3.2 | 25.2 | 243.8 | 264.8 | 102.5 | 8.8 | 4.3 | 0.6 | 0.4 | - | 668.4 |
|  |  |  |  |  | Divisi | Па |  |  |  |  |  |  |
| 1983 | 5.4 | 5.5 | 0.1 | 0.2 | 0.3 | 0.1 | - | - | - | - | 1 | 12.6 |
| 1984 | 4.9 | 14.4 | 5.6 | 0.1 | 0.1 | 0.1 | - | - | - | - | 0.2 | 25.4 |
| 1985 | 3.8 | 7 | 11.7 | 4.1 | 0.1 | - | + | - | - | - | 0.1 | 26.8 |
| 1986 | 0.4 | 0.3 | 3.5 | 10.4 | 2.9 | 0.1 | + | + | - | - | - | 17.6 |
| 1987 | - | - | - | - | 0.3 | 0.3 | - | - | - | - | - | 0.6 |
| 1988 | 1 | 0.1 | - | + | 0.2 | 0.5 | 0.2 | - | - | - | - | 2.1 |
| 1989 | 0.1 | 0.7 | 2.7 | + | 0.1 | 0.1 | 0.1 | - | - | - | - | 3.8 |
| 1990 | 6.1 | 0.9 | 0.9 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | - | - | - | 8.4 |
| 1991 | 5.7 | 3.8 | 0.6 | 0.1 | + | - | - | - | - | - | - | 10.2 |
| 1992 | 1.2 | 2.3 | 5.6 | 2.3 | 3 | 0.3 | 0.3 | 0.4 | 0.4 |  | - | 15.8 |
| 1993 | 1.8 | 1.1 | 1.5 | 4.5 | 2.5 | 0.8 | 0.2 | 0.1 | 0.2 | 0.2 | - | 12.8 |
| 1994 | 1 | 0.6 | 0.5 | 3.1 | 15.9 | 4.4 | 1.5 | + | 0.1 | 0.1 | - | 27.2 |
| 1995 | 5 | 8.5 | 6.3 | 5.3 | 6.2 | 23.9 | 4.1 | 0.6 | + | 0.2 | - | 60.1 |
| 1996 | 29.2 | 4.1 | 25 | 8.1 | 4.9 | 9.1 | 13.4 | 1.3 | 0.4 | 0.1 | - | 95.7 |
| 1997 | 1.2 | 2.8 | 0.8 | 1.3 | 0.7 | 0.6 | 0.9 | 0.5 | 0.1 | - | - | 8.9 |
| 1998 | 23.2 | 7.8 | 15.5 | 1.1 | 2.4 | 3.2 | 0.5 | 2.8 | 0.8 | 0.1 | - | 57.3 |
| 1999 | 34.8 | 34.1 | 4.3 | 16.9 | 3.9 | 6.3 | 1.7 | 0.9 | 1.2 | 0.5 | - | 104.6 |
| 2000 | 27.9 | 23.9 | 13.5 | 1.8 | 9.3 | 2 | 0.9 | 0.2 | 0.2 | 0.4 | - | 80.1 |
| 2001 | 39 | 13.5 | 7.6 | 8.4 | 2.2 | 7.9 | 1.4 | 0.3 | 0.1 | 0.4 | - | 80.8 |
| $2002{ }^{2}$ | 61.9 | 16.6 | 5.3 | 10.2 | 29.9 | 6 | 3.3 | 0.3 | 0.1 | 0.2 | - | 133.7 |
| 2003 | 20.6 | 30.8 | 9.8 | 8.3 | 10.4 | 16.1 | 2.4 | 2.1 | 0.2 | + | - | 100.7 |
| 2004 | 100.2 | 32.8 | 18.1 | 4.5 | 5.5 | 7.2 | 8.1 | 0.7 | 1.1 | 0.3 | - | 178.4 |
| 2005 | 61.6 | 23.9 | 4.6 | 10.9 | 2.1 | 2.7 | 5.3 | 2.9 | 0.5 | 0.2 | - | 114.6 |
| 2006 | 33.3 | 36.9 | 15.2 | 1.9 | 8.2 | 3.4 | 2.5 | 1.8 | 1.8 | 0.3 | - | 105.5 |
| 2007 | 28.2 | 96 | 33.9 | 14.1 | 2.1 | 5.1 | 2.2 | 0.6 | 0.9 | 0.4 | - | 183.4 |
| 2008 | 13.6 | 23.8 | 64.3 | 26.8 | 9.6 | 1.8 | 2.6 | 0.4 | 0.3 | 0.3 | - | 143.6 |
| 2009 | 8.6 | 5.7 | 7.6 | 34.5 | 23.2 | 9.2 | 1.2 | 1.7 | 0.2 | 0.1 | - | 91.9 |

Table B2 (continued)

| Year | Age |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | Total |
| Division IIb |  |  |  |  |  |  |  |  |  |  |  |  |
| 1983 | 22.1 | 9.9 | 0.2 | 0.1 | + | + | - | - | - | - | 0.1 | 32.4 |
| 1984 | 2.2 | 14.3 | 1.8 | - | - | - | - | - | - | - | + | 18.3 |
| 1985 | 1.4 | 10.2 | 61.4 | 5.1 | + | + | + | - | - | - | + | 78.1 |
| 1986 | + | 0.2 | 3.1 | 7.2 | 1.4 | - | + | - | - | - | - | 12 |
| 1987 | - | - | 0.1 | 0.7 | 1.4 | 0.5 | + | - | - | - | - | 2.8 |
| 1988 | 0.2 | - | - | + | 0.3 | 1.1 | 0.2 | - | + | - | - | 1.8 |
| 1989 | 0.7 | 0.1 | 0.2 | + | 0.1 | 0.3 | 0.6 | 0.1 | + | - | - | 2.1 |
| 1990 | 12.9 | 5.4 | 0.8 | + | + | 0.2 | 0.1 | 0.1 | + | - | - | 19.5 |
| 1991 | 20 | 22.9 | 6.2 | 0.4 | 0.1 | 0.1 | 0.1 | + | + | - | - | 49.8 |
| 1992 | 13.3 | 9.1 | 69.8 | 13.9 | 0.5 | + | + | - | + | + | - | 106.6 |
| 1993 | 0.7 | 0.9 | 1.9 | 24.7 | 1.9 | 0.2 | + | + | + | + | - | 30.4 |
| 1994 | 0.4 | 1.7 | 1.7 | 2.3 | 15.7 | 2.7 | 0.8 | 0.2 | + | + | - | 25.5 |
| 1995 | 0.1 | 0.4 | 0.4 | 0.8 | 0.6 | 1.6 | 0.4 | + | + | + | - | 4.3 |
| $1996{ }^{1}$ | 4.3 | 0.6 | 0.5 | 0.3 | 0.2 | 0.4 | 0.5 | 0.3 | - | - | - | 7.1 |
| $1997{ }^{1}$ | 0.4 | 1.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | + | + | - | 2.1 |
| 1998 | 5.8 | 1.1 | 0.2 | + | 0.1 | 0.1 | + | 0.1 | + | - | - | 7.5 |
| 1999 | 8.6 | 20.1 | 1.8 | 1.2 | 0.5 | 0.3 | 0.1 | - | 0.2 | 0.1 | - | 32.9 |
| 2000 | 7.9 | 10 | 13.4 | 1.3 | 5.5 | 2.2 | 1.2 | 0.4 | 0.2 | 0.3 | - | 42.4 |
| 2001 | 2.7 | 13.1 | 15.9 | 11.4 | 0.8 | 4.7 | 1.2 | 0.4 | 0.1 | 0.6 | - | 51 |
| $2002{ }^{2}$ | 9 | 4.2 | 7.7 | 5.1 | 2.6 | 0.7 | 0.8 | 0.1 | 0.1 | 0.1 | - | 30.4 |
| 2003 | 3.6 | 21.5 | 10.4 | 15.5 | 11.3 | 15.9 | 3.6 | 3 | 0.4 | 0.3 | - | 85.7 |
| 2004 | 34.9 | 5.6 | 6.4 | 1.3 | 2.6 | 1.8 | 2.9 | 0.1 | 0.2 | 0.1 | - | 56 |
| 2005 | 60.9 | 43.5 | 4.1 | 10.3 | 4.1 | 2.7 | 3.6 | 2.2 | 0.1 | 0.3 | - | 131.7 |
| $2006{ }^{3}$ | 75.4 | 110.6 | 71.6 | 4.6 | 6.1 | 2.4 | 1.4 | 2 | 1.8 | 0.3 | - | 276.2 |
| 2007 | 3.3 | 67.3 | 396.4 | 78.7 | 5.5 | 26 | 7.3 | 2.9 | 2.6 | 0.8 | - | 590.9 |
| 2008 | 1.5 | 3.8 | 204.1 | 304.3 | 50.7 | 7.4 | 13.6 | 2.9 | 2 | 0.7 | - | 591.9 |
| 2009 | 2.6 | 1.1 | 3.5 | 93.6 | 81 | 22 | 2.4 | 2.1 | 0.3 | 0.5 | - | 209 |
| Total-Sub-area I and Divi sions IIa and IIb |  |  |  |  |  |  |  |  |  |  |  |  |
| 1983 | 29.8 | 59.2 | 9.5 | 0.5 | 0.4 | + | - | - | - | - | 0.8 | 100.2 |
| 1984 | 6.4 | 58.6 | 58.4 | 1.5 | 0.2 | 0.1 | + | - | - | - | 0.3 | 125.5 |
| 1985 | 3 | 14.4 | 134.3 | 90 | 0.4 | 0.1 | 0.1 | - | - | - | 0.2 | 242.7 |
| 1986 | 0.2 | 1.4 | 10.7 | 36.3 | 16.4 | 0.1 | + | + | + | - | + | 65.1 |
| 1987 | 0.3 | 0.9 | 1.7 | 8.3 | 22.5 | 5.7 | + | + | - | + | - | 39.4 |
| 1988 | 1.3 | 0.3 | 0.7 | 1.7 | 4 | 7.6 | 0.8 | + | + | + | - | 16.4 |
| 1989 | 2.2 | 1.8 | 2.4 | 0.4 | 1.4 | 4.1 | 8.1 | 1.1 | 0.1 | + | - | 21.6 |
| 1990 | 44.8 | 14.3 | 10.6 | 7.3 | 4.2 | 7.3 | 7.4 | 5.7 | 0.3 | 0.1 | - | 102 |
| 1991 | 16.7 | 42.9 | 17.6 | 6.2 | 0.9 | 0.3 | 0.6 | 1.8 | 1.5 | 0.2 | - | 88.7 |
| 1992 | 16.4 | 28.2 | 128.6 | 34.6 | 5 | 0.4 | 0.6 | 0.9 | 0.8 | 0.1 | - | 215.6 |
| 1993 | 3.5 | 4.8 | 35.7 | 198.5 | 35.6 | 4.8 | 0.8 | 0.4 | 0.4 | - | - | 284.5 |
| 1994 | 9.1 | 4.9 | 5.8 | 44.2 | 101.4 | 11.6 | 1.5 | 0.1 | 0.1 | 0.5 | - | 179.2 |
| 1995 | 6.4 | 7.2 | 4.2 | 3.1 | 12.3 | 37 | 4 | 0.5 | 0.1 | 0.3 | - | 75.1 |
| $1996{ }^{1}$ | 6 | 2.3 | 5.7 | 2.8 | 4.9 | 36.2 | 33.4 | 2.9 | 0.3 | 0.3 | - | 94.8 |
| $1997{ }^{1}$ | 1.8 | 4.6 | 1.9 | 3.2 | 3.2 | 1 | 2.7 | 1 | 0.8 | - | - | 20.2 |
| 1998 | 10.7 | 2.9 | 11.5 | 3.8 | 4.6 | 0.8 | 0.5 | 1.5 | 0.5 | + | - | 36.8 |
| 1999 | 11.7 | 28.9 | 6.1 | 19.6 | 3.9 | 3.7 | 0.8 | 0.3 | 0.7 | 0.7 | - | 76.4 |
| 2000 | 15.1 | 20.7 | 26.2 | 6 | 10.9 | 2.6 | 1.1 | 0.2 | 0.1 | 0.4 | - | 83.3 |
| 2001 | 20.8 | 14.9 | 26.1 | 33.4 | 4 | 6.5 | 1.1 | 0.4 | 0.1 | 0.3 | - | 107.5 |
| $2002{ }^{2}$ | 33.2 | 19.3 | 18.9 | 39.9 | 45 | 4.7 | 2.4 | 0.4 | 0.1 | 0.2 | - | 164 |
| 2003 | 19.8 | 32.8 | 25.1 | 22.1 | 29.9 | 23.1 | 3.4 | 1.6 | 0.2 | 0.1 | - | 158.3 |
| 2004 | 50 | 11 | 20.6 | 11.3 | 9.4 | 10.7 | 8.7 | 0.5 | 0.4 | 0.2 | - | 122.8 |
| 2005 | 62 | 79.2 | 13.6 | 24 | 8.6 | 4.8 | 5.7 | 2.4 | 0.1 | 0.2 | - | 200.7 |
| $2006{ }^{3}$ | 53.4 | 79.2 | 122.7 | 11.3 | 11.9 | 5.7 | 2.6 | 2.4 | 1.1 | 0.2 | - | 290.5 |
| 2007 | 6.5 | 83.9 | 214.2 | 83.8 | 7.3 | 13.7 | 3.8 | 1.4 | 1.1 | 0.4 | - | 416 |
| 2008 | 5.7 | 12.7 | 232.7 | 255.7 | 105.1 | 12.4 | 11.1 | 1.7 | 0.7 | 0.4 | - | 638.7 |
| 2009 | 10 | 2.9 | 15.8 | 164.7 | 170.4 | 63.1 | 5.7 | 3.2 | 0.5 | 0.4 | - | 436.7 |

${ }^{1}$ Adjusted data based on average 1985-1995 distribution.
${ }^{2}$ Adjusted based on 2001 distribution.
${ }^{3}$ Adjusted based on 2004-2006 distribution.

+ means value <0.1; - means 0 value

Table B3 Northeast Arctic HADDOCK. Results from the Norwegian acoustic survey in the Barents Sea in January-March. Stock numbers in millions. New TS and rock-hopper gear (1981-1988 backcalculated from bobbins gear). Corrected for length dependent effective spread of the trawl.

| Year | Age |  |  |  |  |  |  |  |  |  | Total | Area covered (1000 $\mathrm{nm}^{2}$ ) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |  |  |
| 1981 | 7 | 14 | 5 | 21 | 60 | 18 | 1 | 0 | 0 | 0 | 126 | 88.1 |
| 1982 | 9 | 2 | 3 | 4 | 4 | 10 | 6 | 0 | 0 | 0 | 38 | 88.1 |
| 1983 | 0 | 5 | 2 | 3 | 1 | 1 | 4 | 2 | 0 | 0 | 18 | 88.1 |
| 1984 | 1685 | 173 | 6 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 1867 | 88.1 |
| 1985 | 1530 | 776 | 215 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 2526 | 88.1 |
| 1986 | 556 | 266 | 452 | 189 | 0 | 0 | 0 | 0 | 0 | 0 | 1463 | 88.1 |
| 1987 | 85 | 17 | 49 | 171 | 50 | 0 | 0 | 0 | 0 | 0 | 372 | 88.1 |
| 1988 | 18 | 4 | 8 | 23 | 46 | 7 | 0 | 0 | 0 | 0 | 106 | 88.1 |
| 1989 | 52 | 5 | 6 | 11 | 20 | 21 | 2 | 0 | 0 | 0 | 117 | 88.1 |
| 1990 | 270 | 35 | 3 | 3 | 4 | 7 | 11 | 2 | 0 | 0 | 335 | 88.1 |
| 1991 | 1890 | 252 | 45 | 8 | 3 | 3 | 3 | 6 | 0 | 0 | 2210 | 88.1 |
| 1992 | 1135 | 868 | 134 | 23 | 2 | 0 | 0 | 1 | 2 | 0 | 2165 | 88.1 |
| 1993 | 947 | 626 | 563 | 130 | 13 | 0 | 0 | 0 | 0 | 3 | 2282 | 137.6 |
| 1994 | 562 | 193 | 255 | 631 | 111 | 12 | 0 | 0 | 0 | 0 | 1764 | 143.8 |
| 1995 | 1379 | 285 | 36 | 111 | 387 | 42 | 2 | 0 | 0 | 0 | 2242 | 186.6 |
| 1996 | 249 | 229 | 44 | 31 | 76 | 151 | 8 | 0 | 0 | 0 | 788 | 165.3 |
| $1997{ }^{1}$ | 693 | 24 | 51 | 17 | 12 | 43 | 43 | 2 | 0 | 0 | 885 | 87.5 |
| $1998{ }^{1}$ | 220 | 122 | 20 | 28 | 12 | 5 | 13 | 16 | 1 | 0 | 437 | 99.2 |
| 1999 | 856 | 46 | 57 | 13 | 14 | 4 | 1 | 2 | 2 | 0 | 994 | 118.3 |
| 2000 | 1024 | 509 | 32 | 65 | 19 | 11 | 2 | 1 | 2 | 0 | 1664 | 162.4 |
| 2001 | 976 | 316 | 210 | 23 | 22 | 1 | 1 | 0 | 0 | 1 | 1549 | 164.1 |
| 2002 | 2062 | 282 | 216 | 149 | 14 | 12 | 1 | 0 | 0 | 1 | 2737 | 156.7 |
| 2003 | 2394 | 279 | 145 | 198 | 169 | 17 | 5 | 0 | 0 | 1 | 3208 | 146.6 |
| 2004 | 752 | 474 | 127 | 76 | 76 | 66 | 7 | 2 | 0 | 0 | 1580 | 164.6 |
| 2005 | 3364 | 209 | 219 | 102 | 36 | 40 | 9 | 0 | 0 | 0 | 3979 | 178.9 |
| 2006 | 2767 | 804 | 54 | 86 | 30 | 12 | 9 | 2 | 0 | 0 | 3764 | 1691 |
| $2007{ }^{1}$ | 3197 | 868 | 379 | 54 | 88 | 22 | 6 | 5 | 2 | 0 | 4621 | 122.2 |
| 2008 | 1266.6 | 1835 | 723 | 252 | 57 | 74 | 10 | 6 | 0 | 1 | 4226 | 164.4 |
| 2009 | 849 | 246.3 | 1021.7 | 773 | 402.1 | 31.3 | 14.9 | 1.6 | 0.13 | 0.53 | 3341 | 170.9 |
| 2010 | 2035.8 | 81.8 | 138 | 593 | 557.4 | 191.4 | 10.3 | 2.9 | 0.68 | 0.72 | 3612 | 159.9 |

${ }^{1}$ Indices adjusted to account for limited area coverage.
Survey areas extended from 1993 onwards.

Table B4a. Northeast Arctic HADDOCK. Results from the Russian trawl-acoustic survey in the Barents Sea and adjacent waters in late autumn (old method). Index of number of fish at age (+ means value $<1$; means 0 value).

| Year | Age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ | Total |
| $1985{ }^{1}$ | 194 | 434 | 1468 | 636 | 3 | 1 | + | - | - | 1 | 2737 |
| $1986^{1}$ | 34 | 37 | 208 | 917 | 910 | 2 | + | + | + | + | 2109 |
| $1987{ }^{2}$ | 6 | 16 | 29 | 62 | 197 | 61 | + | - | - | 12 | 383 |
| $1988{ }^{2}$ | 2 | 1 | 3 | 18 | 83 | 301 | 46 | - | - | + | 454 |
| $1989{ }^{1}$ | 41 | 32 | 94 | 2 | 14 | 35 | 67 | 9 | 1 | + | 295 |
| $1990^{1}$ | 594 | 176 | 75 | 28 | 17 | 23 | 43 | 44 | 4 | 1 | 1004 |
| $1991{ }^{1}$ | 240 | 368 | 143 | 65 | 11 | 4 | 7 | 21 | 17 | 2 | 878 |
| $1992{ }^{1}$ | 199 | 245 | 758 | 218 | 35 | 3 | 4 | 7 | 6 | + | 1475 |
| $1993{ }^{1}$ | 20 | 26 | 199 | 1076 | 228 | 31 | 5 | 2 | 3 | 5 | 1595 |
| $1994{ }^{1}$ | 118 | 51 | 39 | 252 | 591 | 76 | 9 | + | 1 | 4 | 1141 |
| $1995{ }^{1}$ | 38 | 40 | 18 | 18 | 77 | 225 | 23 | 3 | 1 | 1 | 443 |
| $1996{ }^{1 / 4}$ | 281 | 44 | 148 | 93 | 69 | 280 | 242 | 19 | 3 | 2 | 1181 |
| $1997{ }^{1,4}$ | 70 | 138 | 41 | 207 | 82 | 48 | 41 | 25 | 20 | - | 671 |
| $1998{ }^{3}$ | 107 | 27 | 82 | 22 | 25 | 7 | 3 | 9 | 3 | + | 284 |
| $1999{ }^{1}$ | 222 | 330 | 43 | 129 | 25 | 29 | 7 | 3 | 7 | 2 | 798 |
| $2000^{1}$ | 246 | 292 | 238 | 49 | 86 | 23 | 9 | 2 | 1 | 4 | 949 |
| $2001^{1}$ | 256 | 122 | 200 | 229 | 24 | 45 | 7 | 3 | 1 | 2 | 888 |
| $2002^{1,5,6}$ | 868 | 811 | 581 | 447 | 237 | 329 | 49 | 20 | 12 | 10 | 3364 |
| $2003{ }^{6}$ | 352 | 310 | 189 | 124 | 161 | 124 | 19 | 9 | 1 | 1 | 1290 |
| 2004 | 3164 | 472 | 421 | 176 | 143 | 154 | 151 | 10 | 21 | 5 | 4722 |
| 2005 | 7156 | 2521 | 271 | 476 | 172 | 114 | 154 | 79 | 5 | 7 | 10956 |
| 2006 | - | - | - | - | - | - | - | - | - | - | - |
| 2007 | - | - | - | - | - | - | - | - | - | - | - |
| 2008 | 106 | 172 | 1960 | 1911 | 783 | 99 | 96 | 15 | 7 | 5 | 5153 |
| 2009 | 302 | 28 | 126 | 943 | 1050 | 445 | 40 | 20 | 3 | 2 | 2959 |

Table B4b. Northeast Arctic HADDOCK. Results from the Russian trawl-acoustic survey in the Barents Sea and adjacent waters in late autumn (ne w method). Index of number of fish at age (+ means value $<1$; means 0 value).

| Year | Age |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | Total |
| $1995{ }^{5}$ | 163 | 170 | 79 | 71 | 230 | 404 | 41 | 5 | 1 | 1 | 2 | 1168 |
| $1996{ }^{1,3}$ | 992 | 245 | 291 | 91 | 63 | 206 | 187 | 17 | 1 | + | + | 2092 |
| $1997{ }^{1,3}$ | 185 | 104 | 21 | 121 | 94 | 48 | 47 | 31 | 20 | + | + | 671 |
| $1998{ }^{2}$ | 257 | 44 | 83 | 20 | 20 | 6 | 2 | 7 | 2 | + | + | 442 |
| $1999{ }^{1}$ | 632 | 499 | 60 | 123 | 14 | 16 | 4 | 1 | 4 | 1 | $+$ | 1355 |
| $2000^{1}$ | 524 | 395 | 287 | 54 | 57 | 14 | 6 | 1 | 1 | 1 | 1 | 1340 |
| $2001^{1}$ | 491 | 160 | 227 | 221 | 19 | 35 | 5 | 2 | 1 | 1 | 1 | 1163 |
| $2002^{1,4,5}$ | 1045 | 209 | 139 | 268 | 239 | 27 | 17 | 2 | 1 | + | 1 | 1947 |
| 2003 | 1168 | 473 | 217 | 116 | 134 | 94 | 14 | 6 | 1 | + | + | 2223 |
| 2004 | 8529 | 1141 | 342 | 116 | 54 | 55 | 44 | 3 | 4 | 1 | 1 | 10289 |
| 2005 | 17782 | 2903 | 123 | 205 | 62 | 33 | 38 | 16 | 1 | 1 | + | 21165 |
| $2006{ }^{6}$ | 9396 | 1286 | 308 | 30 | 31 | 10 | - | 5 | 5 | 4 | 1 | 11075 |
| 2007 | 812 | 1473 | 2226 | 745 | 53 | 75 | 22 | 8 | 7 | 2 | 1 | 5423 |
| 2008 | 245 | 203 | 2134 | 1947 | 728 | 88 | 83 | 13 | 6 | 4 | 2 | 5455 |
| 2009 | 1650 | 204 | 243 | 1455 | 1258 | 485 | 46 | 30 | 4 | 2 | 1 | 5380 |

${ }^{1}$ October-December ${ }^{2}$ September-October ${ }^{3}$ November-January
${ }^{4}$ Adjusted based on average 1985-1995 distribution
${ }^{5}$ Adjusted based on 2001 distribution
${ }^{6}$ Adjusted data in $2004 \quad{ }^{7}$ Not adjusted data to the whole area

Table B5 Northeast Arctic HADDOCK. Length data (cm) from Norwegian surveys in January-March and Russian surveys in November-December.

| Norway |  | Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 |  |  |  |
|  | 1983 | 16.8 | 25.2 | 34.9 | 44.7 | 52.5 | 58.0 | 62.4 |  |  |  |
|  | 1984 | 16.6 | 27.5 | 32.7 | - | 56.6 | 62.4 | 61.8 |  |  |  |
|  | 1985 | 15.7 | 23.9 | 35.6 | 41.9 | 58.5 | 61.9 | 63.9 |  |  |  |
|  | 1986 | 15.1 | 22.4 | 31.5 | 43.0 | 54.6 | - | - |  |  |  |
|  | 1987 | 15.4 | 22.4 | 29.2 | 37.3 | 46.5 | - | - |  |  |  |
|  | 1988 | 13.5 | 24.0 | 28.7 | 34.7 | 41.5 | 47.9 | 54.6 |  |  |  |
|  | 1989 | 16.0 | 23.2 | 31.1 | 36.5 | 41.7 | 46.4 | 52.9 |  |  |  |
|  | 1990 | 15.7 | 24.7 | 32.7 | 43.4 | 46.1 | 50.1 | 52.4 |  |  |  |
|  | 1991 | 16.8 | 24.0 | 35.7 | 44.4 | 52.4 | 54.8 | 55.6 |  |  |  |
|  | 1992 | 15.1 | 23.9 | 33.9 | 45.5 | 53.1 | 59.2 | 60.6 |  |  |  |
|  | 1993 | 14.5 | 21.4 | 31.8 | 42.4 | 50.6 | 56.1 | 59.4 |  |  |  |
|  | 1994 | 14.7 | 21.0 | 29.7 | 38.5 | 47.8 | 54.2 | 56.9 |  |  |  |
|  | 1995 | 15.4 | 20.1 | 28.7 | 34.2 | 42.8 | 51.2 | 55.8 |  |  |  |
|  | 1996 | 15.4 | 21.6 | 28.6 | 37.8 | 42.0 | 46.7 | 55.3 |  |  |  |
|  | 1997 | 16.1 | 27.7 | 27.7 | 35.4 | 39.7 | 47.5 | 50.1 |  |  |  |
|  | 1998 | 14.4 | 29.2 | 29.2 | 35.8 | 41.3 | 48.4 | 50.9 |  |  |  |
|  | 1999 | 14.7 | 20.8 | 32.3 | 39.4 | 45.5 | 52.3 | 54.6 |  |  |  |
|  | 2000 | 15.8 | 22.5 | 30.3 | 41.6 | 47.7 | 50.8 | 51.1 |  |  |  |
|  | 2001 | 22.2 | 22.2 | 32.2 | 37.8 | 47.2 | 51.2 | 58.7 |  |  |  |
|  | 2002 | 21.1 | 21.1 | 29.6 | 40.2 | 44.2 | 50.9 | 58.4 |  |  |  |
|  | 2003 | 16.5 | 24.1 | 28.0 | 37.2 | 46.5 | 49.6 | 54.7 |  |  |  |
|  | 2004 | 14.2 | 22.3 | 30.6 | 36.3 | 43.4 | 49.8 | 51.4 |  |  |  |
|  | 2005 | 15.1 | 20.8 | 30.0 | 36.6 | 41.5 | 47.9 | 51.9 |  |  |  |
|  | 2006 | 14.7 | 22.6 | 31.3 | 37.8 | 43.2 | 48.0 | 50.8 |  |  |  |
|  | $2007{ }^{1}$ | 15.7 | 23.2 | 28.7 | 37.4 | 45.5 | 48.5 | 53.5 |  |  |  |
|  | 2008 | 15.9 | 23.8 | 30.1 | 38.1 | 39.7 | 48.6 | 53.4 |  |  |  |
|  | 2009 | 14.5 | 22.5 | 29.6 | 36 | 41.9 | 46.9 | 51.7 |  |  |  |
|  | 2010 | 14.7 | 20.2 | 30.4 | 37.1 | 41.2 | 45.9 | 50.0 |  |  |  |
| Russia | Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
|  | 1984 | - | 24.1 | 35.8 | 44.4 | 56.4 | 62.8 | 64.8 | - | - | - |
|  | 1985 | 16.5 | 22.4 | 30.9 | 44.1 | 53.8 | 61.3 | 64.7 | - | - | - |
|  | 1986 | 17.0 | 20.7 | 28.1 | 35.4 | 46.7 | 62.0 | - | 68.0 | - | - |
|  | 1987 | 12.1 | 21.5 | 27.8 | 32.3 | 37.3 | 48.6 | - | - | - | - |
|  | 1988 | 13.7 | 23.2 | 29.7 | 33.7 | 39.3 | 46.2 | 51.2 | - | - | - |
|  | 1989 | 14.9 | 22.2 | 26.5 | 38.5 | 44.5 | 49.3 | 53.0 | 57.7 | 64.1 | - |
|  | 1990 | 17.0 | 24.5 | 30.9 | 40.4 | 50.6 | 53.2 | 55.7 | 59.7 | 63.8 | 67.7 |
|  | 1991 | 17.2 | 24.2 | 30.5 | 39.7 | 53.4 | 55.4 | 58.3 | 60.5 | 62.7 | 70.2 |
|  | 1992 | 16.0 | 22.8 | 31.1 | 44.6 | 53.8 | 63.8 | 61.2 | 66.4 | 69.0 | 69.6 |
|  | 1993 | 15.3 | 21.7 | 28.7 | 38.3 | 48.3 | 54.3 | 60.9 | 64.2 | 63.2 | 65.0 |
|  | 1994 | 15.7 | 22.5 | 28.1 | 33.0 | 44.1 | 54.9 | 61.5 | 67.5 | 67.7 | 67.8 |
|  | 1995 | 15.5 | 22.5 | 28.5 | 33.3 | 39.7 | 49.9 | 58.2 | 63.1 | 66.3 | 69.5 |
|  | $1996{ }^{1}$ | 15.8 | 22.8 | 28.4 | 33.7 | 42.0 | 48.7 | 54.8 | 63.4 | 69.3 | 72.0 |
|  | $1997{ }^{1}$ | 13.8 | 23.5 | 29.3 | 36.1 | 45.3 | 50.0 | 54.6 | 58.9 | 69.4 | 66.0 |
|  | 1998 | 15.0 | 22.0 | 29.0 | 38.3 | 47.7 | 52.1 | 54.5 | 57.8 | 63.4 | - |
|  | 1999 | - | 22.8 | 27.4 | 40.1 | 47.4 | 50.9 | 54.6 | 55.9 | 58.0 | 61.6 |
|  | 2000 | 15.0 | 22.7 | 30.4 | 35.2 | 49.3 | 55.1 | 57.8 | 62.4 | 63.3 | 63.6 |
|  | 2001 | 15.1 | 22.4 | 29.8 | 37.8 | 48 | 55.3 | 58.8 | 62.1 | 63.6 | 65.4 |
|  | 2002 | 14.6 | 23.8 | 30.1 | 35.6 | 48.2 | 55.1 | 60.2 | 60.5 | 63.3 | 66.8 |
|  | 2003 | 14.0 | 22.9 | 28.9 | 35.3 | 44.8 | 52.2 | 57.5 | 63.1 | 66.3 | 69.6 |
|  | 2004 | 14.4 | 23.1 | 30.4 | 37.7 | 44.2 | 49.4 | 56.4 | 61.6 | 66.4 | 69.1 |
|  | 2005 | 14.9 | 23.5 | 30.0 | 36.9 | 44.8 | 49.9 | 54.7 | 59.2 | 65.9 | 66.6 |
|  | $2006{ }^{1}$ | 15.3 | 24.1 | 32.6 | 39.8 | 46.7 | 51.8 | 54.9 | 59.0 | 62.4 | 65.3 |
|  | 2007 | 15.4 | 23.7 | 30.6 | 39.2 | 46.6 | 52.0 | 54.4 | 58.4 | 61.3 | 65.8 |
|  | 2008 | 14.5 | 22.3 | 30.8 | 38.1 | 47.3 | 52.8 | 55.8 | 59.1 | 62.8 | 65.0 |
|  | 2009 | 15.4 | 21.8 | 29.4 | 36.0 | 43.9 | 51.0 | 55.3 | 59.2 | 62.3 | 63.3 |

${ }^{1}$ Limited area coverage, lengths are not adjusted to account for limited area coverage.

Table B6 Northeast Arctic HADDOCK. Weight data (g) from Norwegian surveys in January -March and Russian surveys in NovemberDecember.

| Norway | Year /Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1983 | 52 | 133 | 480 | 1043 | 1641 | 2081 | 2592 |  |  |  |  |
|  | 1984 | 36 | 196 | 289 | 964 | 1810 | 2506 | 2240 |  |  |  |  |
|  | 1985 | 35 | 138 | 432 | 731 | 1970 | 2517 | - |  |  |  |  |
|  | 1986 | 47 | 100 | 310 | 734 | - | - | - |  |  |  |  |
|  | 1987 | 24 | 91 | 273 | 542 | 934 | - | - |  |  |  |  |
|  | 1988 | 23 | 139 | 232 | 442 | 743 | 1193 | 1569 |  |  |  |  |
|  | 1989 | 43 | 125 | 309 | 484 | 731 | 1012 | 1399 |  |  |  |  |
|  | 1990 | 34 | 148 | 346 | 854 | 986 | 1295 | 1526 |  |  |  |  |
|  | 1991 | 41 | 138 | 457 | 880 | 1539 | 1726 | 1808 |  |  |  |  |
|  | 1992 | 32 | 136 | 392 | 949 | 1467 | 2060 | 2274 |  |  |  |  |
|  | 1993 | 26 | 93 | 317 | 766 | 1318 | 1805 | 2166 |  |  |  |  |
|  | 1994 | 25 | 86 | 250 | 545 | 1041 | 1569 | 1784 |  |  |  |  |
|  | 1995 | 30 | 71 | 224 | 386 | 765 | 1286 | 1644 |  |  |  |  |
|  | 1996 | 30 | 93 | 220 | 551 | 741 | 1016 | 1782 |  |  |  |  |
|  | 1997 | 35 | 88 | 200 | 429 | 625 | 1063 | 1286 |  |  |  |  |
|  | 1998 | 25 | 112 | 241 | 470 | 746 | 1169 | 1341 |  |  |  |  |
|  | 1999 | 27 | 85 | 333 | 614 | 947 | 1494 | 1616 |  |  |  |  |
|  | 2000 | 32 | 108 | 269 | 720 | 1068 | 1341 | 1430 |  |  |  |  |
|  | 2001 | 28 | 106 | 337 | 556 | 1100 | 1429 | 2085 |  |  |  |  |
|  | 2002 | 30 | 84 | 144 | 623 | 848 | 1341 | 2032 |  |  |  |  |
|  | 2003 | 38 | 127 | 202 | 493 | 981 | 1189 | 1613 |  |  |  |  |
|  | 2004 | 23 | 98 | 266 | 459 | 780 | 1167 | 1328 |  |  |  |  |
|  | 2005 | 29 | 84 | 253 | 469 | 699 | 1054 | 1378 |  |  |  |  |
|  | 2006 | 26 | 107 | 303 | 540 | 821 | 1111 | 1332 |  |  |  |  |
|  | $2007{ }^{1}$ | 32 | 112 | 237 | 539 | 970 | 1195 | 1608 |  |  |  |  |
|  | 2008 | 33 | 115 | 250 | 538 | 692 | 1259 | 1609 |  |  |  |  |
|  | 2009 | 25 | 98 | 230 | 440 | 718 | 1029 | 1402 |  |  |  |  |
|  | 2010 | 28 | 76 | 273 | 473 | 656 | 945 | 1249 |  |  |  |  |
| Russia | Year /Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
|  | 1984 | 36 | 127 | 438 | 815 | 1777 | 2395 | 2688 | - | - | - | - |
|  | 1985 | 37 | 105 | 282 | 817 | 1530 | 2262 | 2263 | - | - | - | - |
|  | 1986 | 38 | 88 | 209 | 419 | 919 | 2240 | - | 3100 | - | - | - |
|  | 1987 | - | 95 | 196 | 330 | 497 | 1055 | - | - | - | - | - |
|  | 1988 | 35 | 106 | 248 | 398 | 627 | 997 | 1431 | - | - | - | - |
|  | 1989 | 52 | 105 | 181 | 606 | 903 | 1287 | 1587 | 2004 | 2716 | - | - |
|  | 1990 | 62 | 143 | 288 | 667 | 1337 | 1533 | 1778 | 2233 | 2731 | 3092 | - |
|  | 1991 | 57 | 133 | 292 | 690 | 1570 | 1863 | 2206 | 2320 | 2568 | 3525 | - |
|  | 1992 | 40 | 108 | 279 | 850 | 1542 | 2199 | 2363 | 3045 | 3391 | 3400 | 4200 |
|  | 1993 | 31 | 96 | 217 | 535 | 1077 | 1493 | 2094 | 2509 | 2374 | 2621 | 3160 |
|  | 1994 | 27 | 106 | 205 | 337 | 841 | 1602 | 2256 | 2913 | 2934 | 3033 | 3163 |
|  | 1995 | 28 | 95 | 196 | 345 | 628 | 1234 | 1908 | 2430 | 2815 | 3323 | 3479 |
|  | 1996 | 30 | 103 | 209 | 347 | 743 | 1152 | 1650 | 2442 | 3218 | 3333 | 4648 |
|  | 1997 | 22 | 115 | 227 | 447 | 911 | 1216 | 1583 | 1966 | 3155 | 2815 | 3423 |
|  | 1998 | 27 | 94 | 230 | 569 | 1087 | 1482 | 1690 | 1914 | 2539 | 3893 | 3900 |
|  | 1999 | - | 104 | 191 | 648 | 1049 | 1251 | 1544 | 1608 | 1814 | 2210 | 2978 |
|  | 2000 | 29 | 110 | 278 | 427 | 1249 | 1681 | 1966 | 2488 | 2625 | 2648 | - |
|  | 2001 | 26 | 102 | 244 | 533 | 1097 | 1695 | 2065 | 2469 | 2704 | 2867 | 3141 |
|  | 2002 | 25 | 127 | 280 | 457 | 1166 | 1690 | 2293 | 2484 | 2784 | 2962 | 4655 |
|  | 2003 | 21 | 104 | 220 | 419 | 855 | 1347 | 1844 | 2402 | 2923 | 2582 | - |
|  | 2004 | 24 | 87 | 253 | 518 | 846 | 1130 | 1571 | 1959 | 2633 | 3366 | - |
|  | 2005 | 27 | 115 | 259 | 511 | 933 | 1289 | 1670 | 2079 | 2833 | 2965 | - |
|  | $2006{ }^{1}$ | 26 | 105 | 269 | 444 | 867 | 1307 | 1604 | 1922 | 2274 | 2520 | - |
|  | 2007 | 30 | 117 | 274 | 600 | 1012 | 1436 | 1647 | 2018 | 3214 | 2885 | - |
|  | 2008 | 25 | 94 | 267 | 545 | 1046 | 1445 | 1755 | 2126 | 2458 | 2735 | 3289 |
|  | 2009 | 28 | 91 | 241 | 448 | 841 | 1335 | 1666 | 2048 | 2438 | 2498 | 3132 |

${ }^{1}$ Limited area coverage, weights are not adjusted to account for limited area coverage.

## 5 Saithe in Subareas I and II (Northeast Arctic)

An update assessment is presented for this stock. The last benchmark assessment was done at WKROUND February 2010 (ICES CM 2010/ACOM:36). The main conclusions of the benchmark assessment were:

- Expand the catch matrix from 3-11+ to 3-15+
- Base the Norwegian trawl CPUE on data from all quarters and from days with $>20 \%$ but $<80 \%$ saithe in the catches
- Split the two tuning series in 2002
- Reduce the shrinkage in the XSA and remove the time tapered downweighting

More details and general information is given in (ICES CM 2010/ACOM:36) and the Stock Annex (Quality Handbook).

### 5.1 The Fishery (Tables 5.1.1-5.1.2, Figure 5.1.1)

Currently the main fleets targeting saithe include trawl, purse seine, gillnet, hand line and Danish seine. Landings of saithe were highest in 1970-1976 with an average of $239,000 \mathrm{t}$ and a maximum of $265,000 \mathrm{t}$ in 1974. This period was followed by a sharp decline to a level of about 160,000 $t$ in the years 1978-1984. Another decline followed and from 1985 to 1991 the landings ranged from 67,000-123,000 t. After 1990 landings increased again and reached $171,000 \mathrm{t}$ in 1996, followed by a new decline to $136,000 \mathrm{t}$ in 2000 and 2001. Since then the annual landings have increased gradually to 212,000 t in 2006, followed by a decline to 199000 t in $2007,183,000 \mathrm{t}$ in 2008 and $161,000 \mathrm{t}$ in 2009.

Discarding, although illegal, occurs in the saithe fishery, but is not considered a major problem in the assessment. Due to its near-shore distribution saithe is virtually inaccessible for commercial gears during the first couple of years of life and there are no reports indicating overall high discard rates in the Norwegian fisheries. There are reported incidents of slipping in the purse seine fishery, mainly related to minimum landing size. On trawlers, discarding may occur when vessels targeting other species catch saithe, for which they may not have a quota or have filled it, and there are undocumented observations and comparisons of scientific samples from nonNorwegian commercial trawlers indicating that discarding may be substantial in certain areas and seasons. However, there are no quantitative estimates of the level of discarding available.

### 5.1.1 ICES advice applicable to 2009 and 2010

The advice from ICES for 2009 was as follows:
Exploitation boundaries in relation to proposed and evaluated management plan: The implemented management plan implies a TAC based on the average catches for the coming 3 years based on $\mathrm{F}_{\mathrm{pa}}$. This results in a TAC of 225000 t in 2009, and a fishing mortality of 0.29 .

Exploitation boundaries in relation to high long-term yield, low risk of depletion of production potential, and considering ecosystem effects: The current fishing mortality is lower than the F associated with high long-term yield when applied within the agreed HCR.

Exploitation boundaries in relation to precautionary limits: The implemented management plan has been found to be consistent with the precautionary approach and ICES therefore advises according to this plan. This results in a TAC of 225000 t in 2009.

The advice from ICES for 2010 was as follows:
Exploitation boundaries in relation to proposed and evaluated management plan: The implemented management plan implies a TAC based on the average catches for the coming 3 years based on $F_{p a}$. This results in a TAC of 204000 t in 2010, and a fishing mortality of 0.30 .

Exploitation boundaries in relation to high long-term yield, low risk of depletion of production potential, and considering ecosystem effects: The current fishing mortality is lower than the F associated with high long-term yield when applied within the agreed HCR.

Exploitation boundaries in relation to precautionary limits: The implemented management plan has been found to be consistent with the precautionary approach and ICES therefore advises according to this plan. This results in a TAC of 204000 t in 2010.

### 5.1.2 Management applicable in 2009 and 2010

Management of Saithe in Sub-areas I and II is by TAC and technical measures. Norwegian authorities set the TACs for 2009 and 2010 to $225,000 \mathrm{t}$ and 204,000 t, respectively. The Institute of Marine Research, Bergen, Norway (IMR), advised a TAC for 2009 of $214,000 \mathrm{t}$, estimated by applying a fishing mortality of 0.32 to the HCR, i.e. a little below the target F of $0.35\left(\mathrm{~F}_{\mathrm{pa}}\right)$ specified in the HCR . Following the same principle, IMR advised a TAC of 193000 t for 2010. ICES, in the evaluation of the management plan, also recommended using 0.32 , corresponding to the highest long-term yield, in the HCR (ICES Advice 2007).

### 5.1.3 The fishery in 2009 and expected landings in 2010

Provisional figures show that the landings in 2009 were approximately $161,000 \mathrm{t}$, which is about $64,000 \mathrm{t}$ less than the TAC of $225,000 \mathrm{t}$, which also were expected landings in the forecast last year.

Official landings in 2010 will probably also be less than the TAC of 204,000 $t$, which is only 9 \% less than the 2009 TAC, but 26 \% higher than the 2009 landings. However, since the WG does not have any prognosis of total landings in 2010 available, the TAC of $204,000 \mathrm{t}$ is used in the projections.

### 5.2 Commercial catch-effort data and research vessel surveys

### 5.2.1 Fishing Effort and Catch-per-unit-effort (Tables 5.2.1)

In the Norwegian trawl CPUE indices, all quarters and all days with more than 20 \% but less than $80 \%$ saithe in the catches from vessels larger than the median length were included. The $80 \%$ limit was set to get a more consistent time series regarding bycatch or direct saithe fishery (Fotland et al., WD 12 WKROUND 2010). Since the 2007 WG double and triple trawl catches have been excluded from the data because such trawls have a much higher efficiency and the use of them have increased over the last few years. The CPUE observations were averaged over each quarter, and then a yearly index was calculated by averaging over the year. The total CPUE index was finally divided on age groups applying yearly catch in numbers and weight at age data from the trawl fishery.

### 5.2.2 Survey results (Table 5.2.2, Figure 5.2.1)

In autumn 2003 the saithe and coastal cod surveys were combined (Berg et al.,WD11 2004). Exploratory runs with new tuning time series from the combined survey were prepared to the benchmark assessment 2010 (Mehl and Fotland, WD 8 WKROUND 2010). The XSA diagnostics and results showed that the new tuning series did not perform as well as the one presently used. The new ones are still too short for tuning of the XSA and the old one will be applied. The estimation of abundance indices is as far as possible done as before the combination of the two surveys. The total index for 2009 (Mehl et al., WD 8) increased by almost $40 \%$ compared to 2008, but is still one of the lowest since 1991. All age groups are all below the 1992-2008 average. In recent years the proportion of saithe in the southern part of the survey area (sub areas C+D) has increased, from about $30 \%$ in 1997-2002 to $60 \%$ in later years (Figure5.2.1).

### 5.2.3 Recruitment indices

Owing to the near-shore distribution of juvenile saithe, obtaining early estimates of recruitment is a common problem in saithe stocks. Attempts at establishing year class strength at ages 0-2 for the Northeast Arctic saithe stock have so far failed. The survey recruitment indices are strongly dependent on the extent to which 2-4 year old saithe have migrated from the coastal areas and become available to the acoustic saithe survey on the banks, and this varies between years. An observer programme for establishing an 0-group index series started in 2000 (Borge and Mehl, WD 21 2002). However, these observations do not seem to reflect the dynamics in year class strength very well and are probably not suitable for improving future recruitment estimates for this stock (Mehl, WD 6 2007; Mehl, WD 7 to WKROUND 2010). It is therefore decided to terminate the programme.

### 5.3 Data used in the Assessment

### 5.3.1 Catch numbers at age (Tables 5.3.1-5.3.2)

Landings data, logbook adjusted for trawl, and allocation of biological samples of catch numbers, mean length and mean weight at age from the Norwegian fishery in 2008 was updated applying the same method as previously used. The new allocation resulted in a much higher number of 3-year olds in the purse seine catches in 2008. Also for 2007 and 2006 all Norwegian landings data were updated and catch numbers and weights at age were recalculated. For all countries the landings data for 19732008 were updated to the official total catch reported to ICES or to Norwegian authorities. The total landings by numbers for the whole time series back to 1960 were expanded to $15+$, adjusted to the official total catch reported to ICES (Fotland and Mehl, WD 5). These revisions resulted in only minor changes in catch numbers-at-age and weight-at-age.

Age composition data for 2009 were available from Norway, Russia (Sub-areas I and II) and Germany (Subarea II). These countries accounted for $98 \%$ of the landings. Other areas and countries were assumed to have the same age composition as Norwegian trawlers. Table 5.3.1 presents the Norwegian sampling level in 2009. The biological sampling of some vessel groups may have become critically low after the termination of the Norwegian port sampling program in 2009. The 2008 and 2009 catch and sample data were uploaded to the InterCatch database, and there were only minor discrepancies between data allocated and aggregated in InterCatch and data from the spreadsheets used until now (see Section 0.9).

### 5.3.2 Weight at age (Table 5.3.3)

Constant weights at age values are used for the period 1960-1979. For subsequent years, annual estimates of weight at age in the catches are used. Weight at age in the stock is assumed to be the same as weight at age in the catch. Compared to the previous years, there were only small differences in weight at age for the most important age groups in 2009.

### 5.3.3 Natural mortality

A fixed natural mortality of 0.2 for all age groups was used both in the assessment and the forecast.

### 5.3.4 Maturity at age (Table 5.3.4)

A constant maturity ogive was used until the 2005 WG, when these estimates were evaluated. In later years the maturity at age had decreased somewhat, and the WG decided to use a 3-year running average for the period from 1985 and onwards (2year average for the first and last year). New analyses were only available back to 1985. Table5.3.3 presents the 3-year running average maturity ogive.

### 5.3.5 Tuning data (Table 5.3.5)

Until the 2005 WG, the tuning was based on three data series: CPUE from Norwegian purse seine and Norwegian trawl and indices from a Norwegian acoustic survey. The 2005 WG found rather large and variable log q residuals and large S.E. $\log q$ for the purse seine fleet, as well as strong year effects, and in the combined tuning the fleet got low-scaled weights. The WG decided not to include the purse seine tuning fleet in the analysis. This was confirmed by new analyses at the 2010 benchmark assessment (ICES CM 2010/ACOM:36).

Analyses of the two remaining tuning series done at the 2010 benchmark assessment indicated that there had been a shift in catchability around year 2002. The survey was redesigned in 2003, and the fishery to a larger degree targeted older ages. Permanent breaks were made in both tuning series in 2002. The following four tuning fleets are used in the present assessment:

Fleet 11: CPUE data from the Norwegian trawl fisheries 1994-2001, age groups 4 to 8, quarter 1-4.

Fleet 12: CPUE data from the Norwegian trawl fisheries 2002-2009, age groups 4 to 8, quarter 1-4.

Fleet 13: Indices from the Norwegian acoustic survey 1994-2001, age groups 3 to 7.

Fleet 14: Indices from the Norw egian acoustic survey 2002-2009, age groups 3 to 7.

### 5.4 Exploratory runs (Table 5.4.1, Figure 5.4.1)

The settings of the different runs are shown in Table 5.4.1 and the results are given in Figures 5.4.1. The recommendation from the benchmark assessment in 2010 (ICES CM 2010/ACOM:36) was to run the XSA with a 15+ catch matrix, tuning time series broken in 2002, reduced shrinkage (S.E. of the mean to which the estimate are shrunk increased from 0.5 to 1.5 ) and no tapered time w eighting.

Based on the update of catch statistics and allocation of biological samples, a SPALY (Same Procedure As Last Year) XSA (run 1) was performed, giving somewhat different results compared to the 2009 assessment. F4-7 in 2008 is estimated to 0.20 in both runs, while SSB in 2008 decreased from $776,000 \mathrm{t}$ to $751,000 \mathrm{t}$ (Figure5.4.1). Due to the reallocation of biological samples for 2008, resulting in higher landings of three year olds in the purse seine fishery, the estimated number of recruits at age three in 2008 increased from about 250 millions to 500 millions in the SPALY run.

Two single fleet tuning runs were performed; one with the Norwegian trawl CPUE (run 2) and one with the Norwegian acoustic survey (run 3). The last run (4) was with combined fleets.

Figure 5.4.1, in addition to the 2009 update, also compares estimates of SSB and $\mathrm{F}_{4-7}$ in 2009 from the two single fleet XSA -runs and the combined tuning runs. Due to the expansion of the age span from $11+$ to $15+$ and the changes made to the XSA parameter settings (ICES CM 2010/ACOM:36), the 2009 assessment have much lower $\mathrm{F}_{4-7}$ and higher SSB than the new assessment with 2009 as the last data year. The single fleet tuning run based on the CPUE give the lowest $\mathrm{F}_{4-7}$ and highest SSB in the last assessment year (2009), while the run based on the acoustic indices gave similar SSB but considerable higher $\mathrm{F}_{4-7}$ ( 0.29 compared to 0.20 ). The combined run gave the lowest SSB and a slightly higher $\mathrm{F}_{47}$ than the acoustic single fleet run. This run was used as the final run. Compared to the corresponding run made at the benchmark assessment, $\mathrm{F}_{47}$ in 2008 is somewhat higher and SSB lower, mainly due to one additional year of data (2009). The run made at WKROUND only had data back to 1989, and the runs are therefore not directly comparable.

### 5.5 Final assessment run (Tables 5.5.1-5.5.7, Figures 5.5.1-5.5.4)

Extended Survivors Analysis (XSA) was used for the final assessment with settings shown in Table 5.4.1. The settings are in accordance with the recommendations from the benchmark assessment in February 2010 (ICES CM 2010/ACOM:36). Full tuning fleet diagnostics are given in Table 5.5.1.

Figure 5.5.1 presents $\log \mathrm{q}$ residuals for the tuning fleets with the two parts combined. There are some year and age effects in both fleets, especially for the CPUE series. Figure 5.5 .2 presents S.E. $\log \mathrm{q}$ for the different age groups in the fleets used for tuning. The two oldest tuning series have higher S.E. log q, except for age 4 of the latest trawl CPUE series. Figure 5.5 .3 shows estimates of survivors from different fleets and shrinkage, as well as their different weighting in the final XSA-run. The survey gets the highest weights for age groups 3-6. Figure 5.5.4a-b shows plots of the tuning indices versus stock numbers from the XSA.

### 5.5.1 Fishing mortalities and VPA (Tables 5.5.2-5.5.7, 5.7.1, Figure 5.5.5)

The fishing mortality ( $\mathrm{F}_{47}$ ) in 2008 was 0.25 , which is higher than the value of 0.20 from last year's assessment. The main reason for this is the above mentioned changes made to the assessment. The fishing mortality ( $\mathrm{F}_{47}$ ) in 2009 was 0.27 , i.e. slightly above the corresponding figure for 2008 and below the $\mathrm{F}_{\mathrm{pa}}$ of 0.35 . Fishing mortality and stock size have in the last decade been over- and underestimated, respectively, in the last assessment year. Due to the changes made to the assessment, the retrospective pattern has improved considerably, as is illustrated in Figure 5.5.5.

The XSA-estimates of the 2006-2007 year classes are not considered to be reliable and are therefore shaded (Tables 5.5.3 and 5.5.5). In the projections, both were set to the long-term geometrical mean, the value of the 2005-year class at age 4 being obtained
by applying Pope's approximation. The figures are given in input data for prediction (Table 5.7.1). The 2002 year class was the most numerous in the landings for several years and is estimated to be of the same strength as the very strong 1989 and 1992year classes. The 2003-year class is confirmed to be one of the weakest in the time series, and the 2004-year class is also poor, while the 2005-year class seems to be slightly above average strength. Little information is available on the strength of recent year classes.

The total biomass (ages 3+) has been above the long-term (1960-2008) mean since 1995 , reached a maximum in 2005, and is presently declining. The SSB has been above the long-term mean since 2001 and above $B_{\text {pa }}$ since 1995 (Tables 5.5.5-5.5.7). It has declined since 2005, but is still estimated to be twice the $B_{p a}$.

### 5.5.2 Recruitment (Table 5.3.1, Figure 5.1.1)

Estimates of the recruiting year classes up to the 2005-year class (4 year olds) from the XSA were accepted. Catches of age group 3 were low in 2006 and 2007, increased considerably in 2008 and decreased again in 2009 (Table 5.3.1). Until the 2005 WG, RCT3-runs were conducted to estimate the corresponding year classes, with 2 and 3 year olds from the acoustic survey as input together with VPA numbers. These estimates were, however, strongly weighted towards the mean value of the input XSAnumbers, which due to the short survey time series also contained year classes that were still not converged. It has therefore been stated several times in the ACOM Technical Minutes that it would be more transparent to use the long-term GM (geometric mean) recruitment.

The GM recruitment 1960-2008 is 169 million 3 year olds, and this value is used for the 2006-year class. The value is lower than the GM recruitment 1995-2008 ( 181 million), a period where the SSB has been above $\mathrm{B}_{\mathrm{pa}}$.

### 5.6 Reference points (Figure 5.6.1)

In 2010 the age span was expanded from 11+ to 15+ and important XSA parameter settings were changed (ICES CM 2010/ACOM:36). This resulted in changes in estimated fishing mortality, spawning stock biomass and recruitment, especially in the last part of the time series (Figure 5.6.1). Therefore the LIM and PA reference points were re-estimated at the 2010 WG according to the methodology outlined in ICES CM 2003/ACFM:15, while the PA reference point estimation was based on the old procedure (ICES CM 1998/ACFM:10). The results of the segmented regression were not very much different from the previous analyses. The HCR is based on the PA reference points, and if new ones are introduced, the HCR would have to be evaluated again. Due to lack of time to do this during the WG and the transition to MSY based reference points (see Section 0.10), it was decided to not change the existing LIM and PA reference points. The estimations done at the present $W G$ are, however, presented below. No attempts were made to set MSY reference points (Fmsy and MSY Btrigger), see Section 0.10.

### 5.6.1 Biomass reference points

Parameter values, including the change-point, were computed using segmented regression on the 1960-2005 time series of SSB-recruitment pairs. The maximum likelihood estimate of the spawning stock biomass at which recruitment is impaired was $118,542 \mathrm{t}$. Applying the "magic formula" $\mathrm{B}_{\mathrm{pa}}=\mathrm{B}_{\lim } \exp \left(1.645^{*} \sigma\right)$, with a value of 0.3 for
$\sigma$, gave a $\mathrm{B}_{\mathrm{pa}}$ of $194,176 \mathrm{t}$. However, as explained above, the existing values of $\mathrm{B}_{\mathrm{lim}}=$ $136,000 t$ and $B_{p a}=220,000 \mathrm{t}$ will still be used.

### 5.6.2 Fishing mortality reference points (Tables 5.6.1, 5.7.1, Figure 5.1.1)

Flim was set on the basis of $B_{\lim }$ (ICES CM 2003/ACFM:15). The functional relationship between spawner-per-recruit and F gave the F associated with the R/SSB slope derived from the $B \lim$ estimate obtained from the segmented regression. Arithmetic means of proportion mature 1960-2009, w eight in stock and weight in catch 1980-2009 (weights were constant before 1980), natural mortality and fishing pattern 1960-2009 were used for calculating the spawner-per-recruit function using ICES Secretariat yield-per-recruit software. $\mathrm{R} / \mathrm{SSB}=1.48$ from the Blim estimation gave $\mathrm{SSB} / \mathrm{R}=0.676$ and a $F_{l i m}=0.59$. Applying the "magic formula" $F_{p a}=F_{\lim } \exp \left(-1.645^{*} \sigma\right)$, gave a $F_{p a}$ of 0.36. As explained above, the existing values of $\mathrm{F}_{\mathrm{lim}}=0.58$ and $\mathrm{F}_{\mathrm{pa}}=0.35$ will still be used.

Yield and SSB per recruit were based on the parameters in Table 5.7.1 and are presented in Table 5.6.1. and $\mathrm{F}_{35 \% \text { SPR }}$ w ere estimated to be $0.08,0.33$ and 0.11 , respectively. $\mathrm{F}_{0.1}, \mathrm{~F}_{\max }$ have decreased from last year's estimates of 0.16 and 0.39 , respectively. The plot of SSB versus recruitment is shown in Figure 5.1.1. These points are Fmsy candidates, but the estimates, especially of $\mathrm{F}_{\text {max }}$, are unstable for this stock. When the HCR was evaluated (see below), the highest long-term yield was obtained for an exploitation level of 0.32.

### 5.6.3 Harvest control rule

In 2007 Norway asked ICES to evaluate whether a proposal for a harvest control rule for setting the annual fishing quota (TAC) for Northeast Arctic saithe was consistent with the precautionary approach. The harvest control rule contains the following elements:

- Estimate the average TAC level for the coming 3 years based on $\mathbf{F}_{\text {pa. }}$ TAC for the next year will be set to this level as a starting value for the 3-year period.
- The year after, the TAC calculation for the next 3 years is repeated based on the updated information about the stock development. However, the TAC should not be changed by more than $15 \%$ compared with the previous year's TAC.
- If the spawning stock biomass (SSB) in the beginning of the year for which the quota is set (first year of prediction), is below $\mathbf{B}_{\mathrm{pa}}$, the procedure for establishing TAC should be based on a fishing mortality that is linearly reduced from $\mathbf{F}_{\mathrm{pa}}$ at $\mathrm{SSB}=\mathbf{B}_{\mathrm{pa}}$ to 0 at SSB equal to zero. At SSB levels below $\mathbf{B}_{\mathrm{pa}}$ in any of the operational years (current year and 3 years of prediction) there should be no limitations on the year-to-year variations in TAC.

ICES concluded that the HCR is consistent with the precautionary approach for all simulated data and settings, including a rebuilding situation under the condition that the assessment uncertainty and error are not greater than those calculated from historic data. This also holds true when an implementation error (difference between TAC and catch) equal to the historic level of $3 \%$ is included.

The highest long-term yield was obtained for an exploitation level of 0.32, i.e. a little below the target F used in the $\mathrm{HCR}\left(\mathrm{F}_{\mathrm{pa}}\right)$, and ICES recommended using a lower value in the HCR.

The HCR is expected to rebuild a depleted stock to a level above Blim within three years.

### 5.7 Predictions

### 5.7.1 Input data (Table 5.7.1)

The input data to the predictions based on results from the final XSA are given in Table 5.7.1. The stock number at age in 2010 was taken from the XSA for age 5 (2005 year class) and older. The recruitment at age 3 in the last assessment year (2009) was calculated as the long-term GM (geometric mean) recruitment 1960-2008 (Section 5.5.2), and the corresponding numbers at age 4 in the intermediate year (2010) was calculated applying a natural mortality of 0.2 and using Pope's approximation (as recommended by the ACOM reviewers in 2008). The GM age 3 recruitment of 169 million was also used for the 2007 and subsequent year classes. The natural mortality of 0.2 is the same as used in the assessment. For exploitation pattern the average of 2007-2009 was used for age groups 3-10, while for age groups 11-15+ the 2007-2009 average for ages 11-13 was applied for all ages. For weight at age in stock and catch the average of the last three years in the XSA was used. For maturity at age the average of the 2008-2009 annual determinations was applied.

### 5.7.2 Catch options for 2011 (short-term predictions) (Tables 5.7.2-5.7.4)

The management option table (Table 5.7.2) shows that the expected catch of $204,000 \mathrm{t}$ in 2010 will increase the fishing mortality compared to 2009 from 0.27 to 0.32 , which is below the $\mathrm{F}_{\mathrm{pa}}$ of 0.35 . A catch in 2011 corresponding to the $\mathrm{F}_{\text {staus quo }}$ level ( 3 -year average 2007-2009) of 0.25 will be $143,000 t$, while a catch in 2011 corresponding to the evaluated and implemented HCR (average TAC level for the coming 3 years based on $\mathrm{F}_{\mathrm{pa}}$, see Table 5.7.3) is $173,000 \mathrm{t}$. According to the HCR the TAC should not be change by more than $15 \%$ compared with the previous year's TAC as long as SSB is above the $B_{p a}$ of 220,000 , corresponding to a minimum TAC of $173,400 t$ in 2011. This catch corresponds to a fishing mortality of 0.31 in 2011.

For a catch in 2010 corresponding to the HCR, i.e. 204,000 $t$, the SSB is expected to decrease from about $416,000 \mathrm{t}$ at the beginning of 2010 to $357,000 \mathrm{t}$ at the beginning of 2011. At $\mathrm{F}_{\text {status quo in }} 2011 \mathrm{SSB}$ is estimated to decrease to $350,000 \mathrm{t}$ at the beginning of 2012 and for a catch corresponding to the HCR it will decrease to about $324,000 \mathrm{t}$. Higher fishing mortalities and incoming year classes of below average strength mainly explain this predicted reduction in SSB. Table 5.7.4 presents detailed output for fishing according to the HCR in 2011.

### 5.7.3 Medium term simulations (Figure 5.7.1a-b)

The ACOM review groups have not considered the medium term analyses reliable as the results are mainly driven by the assumption of mean recruitment and ignoring the bias in the assessment. Although the recent assessment indicates a reduction of the bias problem, no improved recruitment estimates are available. How ever, the WG made medium-term simulations just to illustrate a scenario following the HCR.

The input data were the same as used for the short-term predictions (Table 5.7.1). Following the HCR, the catch will decrease to $150,000 \mathrm{t}$ in 2014, while the SSB will be reduced to $290,000 \mathrm{t}$. The highest long term yield for GM recruitment and present exploitation pattern and weight at age is about $140,000 \mathrm{t}$ (Table 5.6.1).

### 5.7.4 Comparison of the present and last year's assessment

The current assessment estimated the total stock in 2009 to be $19 \%$ lower and the SSB $34 \%$ lower, compared to the previous assessment. The F in 2008 is estimated to be higher than in the previous assessment and the realized F in 2009 is comparable to the predicted one based on the TAC.

|  | TOTAL STOCK (3+) <br> BY 1 JANUARY <br> 2009 <br> (TONNES) | SSB BY 1 JANUARY <br> 2009 <br> (TONNES) | $\mathrm{F}_{4-7}$ IN 2009 | $\mathrm{~F}_{477}$ IN 2008 |
| :--- | :--- | :--- | :--- | :--- |
| WG 2009 (11+) | 1012184 | 689583 | 0.28 (TAC <br> constraint) | 0.20 |
| WG 2010 (15+) | 798292 | 456509 | 0.27 | 0.25 |

### 5.8 Comments on the assessment and the forecast (Figures 5.8.1a-b, 5.8.2).

The retrospective pattern has been a major concern in the assessment, but due to the changes done at the benchmark assessment (ICES CM 2010/ACOM:36), the assessment has become more stable. The tendency to overestimate F and underestimate SSB in the last assessment year seems to have changed to an opposite situation, but the differences are less than in previous assessments.
Lack of reliable recruitment estimates is still a major problem. Prediction of catches beyond the TAC year will, to a large extent, be dependent on assumptions of average recruitment.

### 5.9 Response to ACOM technical minutes

The major comments made by the five last reviews were handled with during the benchmark assessment in February 2010 (ICES CM 2010/ACOM:36).

The 2009 reviewers commented that the level of discarding (by age) might have some impact on the perception of the stock dynamics (recruitment). If saithe age 3 is important component of discarding, than omitting it in the assessment gives an underestimated recruitment level. This potential problem has not been looked into.

Saithe has recently been more distributed southward and such was the biological sampling activity for estimating maturity ogives. Higher maturity rate in the southern area is observed. The 3 -year running average ogive used in the assessment is not weighted by abundance and in consequence it probably results in biased estimate of maturity ogive in the context of the whole stock. This problem has neither been resolved.

Table 5.1.1 Saithe in Sub-areas I and II (Northeast Arctic).

| Year | Faroe Islands | France | Germany Dem.Rep | Fed.Rep. Germany | Iceland | Norway | Poland | Port ugal | Russia3 | Spain | UK | $\begin{aligned} & \hline \text { Oth } \\ & \text { ers } \\ & 5 \\ & \hline \end{aligned}$ | Total all countries |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | 23 | 1700 |  | 25948 |  | 96050 |  |  |  |  | 9780 | 14 | 133515 |
| 1961 | 61 | 3625 |  | 19757 |  | 77875 |  |  |  |  | 4595 | 18 | 105951 |
| 1962 | 2 | 544 |  | 12651 |  | 101895 |  |  | 912 |  | 4699 | 4 | 120707 |
| 1963 |  | 1110 |  | 8108 |  | 135297 |  |  |  |  | 4112 |  | 148627 |
| 1964 |  | 1525 |  | 4420 |  | 184700 |  |  | 84 |  | 6511 | 186 | 197426 |
| 1965 |  | 1618 |  | 11387 |  | 165531 |  |  | 137 |  | 6741 | 181 | 185600 |
| 1966 |  | 2987 | 813 | 11269 |  | 175037 |  |  | 563 |  | 13078 | 41 | 203788 |
| 1967 |  | 9472 | 304 | 11822 |  | 150860 |  |  | 441 |  | 8379 | 48 | 181326 |
| 1968 |  |  | 70 | 4753 |  | 96641 |  |  |  |  | 8781 |  | 110247 |
| 1969 | 20 | 193 | 6744 | 4355 |  | 115140 |  |  |  |  | 13585 | 23 | 140060 |
| 1970 | 1097 |  | 29362 | 23466 |  | 151759 |  |  | 43550 |  | 15469 |  | 264924 |
| 1971 | 215 | 14536 | 16840 | 12204 |  | 128499 | 6017 |  | 39397 | 13097 | 10361 |  | 241272 |
| 1972 | 109 | 14519 | 7474 | 24595 |  | 143775 | 1111 |  | 1278 | 13125 | 8223 |  | 214334 |
| 1973 | 7 | 11320 | 12015 | 30338 |  | 148789 | 23 |  | 2411 | 2115 | 6841 |  | 213859 |
| 1974 | 46 | 7119 | 29466 | 33155 |  | 152699 | 2521 |  | 28931 | 7075 | 3104 | 5 | 264121 |
| 1975 | 28 | 3156 | 28517 | 41260 |  | 122598 | 3860 | 6430 | 13389 | 11397 | 2763 | 55 | 233453 |
| 1976 | 20 | 5609 | 10266 | 49056 |  | 131675 | 3164 | 7233 | 9013 | 21661 | 4724 | 65 | 242486 |
| 1977 | 270 | 5658 | 7164 | 19985 |  | 139705 | 1 | 783 | 989 | 1327 | 6935 |  | 182817 |
| 1978 | 809 | 4345 | 6484 | 19190 |  | 121069 | 35 | 203 | 381 | 121 | 2827 |  | 155464 |
| 1979 | 1117 | 2601 | 2435 | 15323 |  | 141346 |  |  | 3 | 685 | 1170 |  | 164680 |
| 1980 | 532 | 1016 |  | 12511 |  | 128878 |  |  | 43 | 780 | 794 |  | 144554 |
| 1981 | 236 | 218 |  | 8431 |  | 166139 |  |  | 121 |  | 395 |  | 175540 |
| 1982 | 339 | 82 |  | 7224 |  | 159643 |  |  | 14 |  | 732 |  | 168034 |
| 1983 | 539 | 418 |  | 4933 |  | 149556 |  |  | 206 | 33 | 1251 |  | 156936 |
| 1984 | 503 | 431 | 6 | 4532 |  | 152818 |  |  | 161 |  | 335 |  | 158786 |
| 1985 | 490 | 657 | 11 | 1873 |  | 103899 |  |  | 51 |  | 202 |  | 107183 |
| 1986 | 426 | 308 |  | 3470 |  | 63090 |  |  | 27 |  | 75 |  | 67396 |
| 1987 | 712 | 576 |  | 4909 |  | 85710 |  |  | 426 |  | 57 | 1 | 92391 |
| 1988 | 441 | 411 |  | 4574 |  | 108244 |  |  | 130 |  | 442 |  | 114242 |
| 1989 | 388 | $460{ }^{2}$ |  | 606 |  | 119625 |  |  | 506 | 506 | 726 |  | 122817 |
| 1990 | 1207 | $340{ }^{2}$ |  | 1143 |  | 92397 |  |  | 52 |  | 709 |  | 95848 |
| 1991 | 963 |  | 2reenland | 2003 |  | 103283 |  |  | $504{ }^{4}$ |  | 492 | 5 | 107327 |
| 1992 | 165 | 1980 | 734 | 3451 |  | 119763 |  |  | 964 | 6 | 541 |  | 127604 |
| 1993 | 31 | 566 | 78 | 3687 | 3 | 140604 |  | 1 | 9509 | $4^{2}$ | 415 | $5^{2}$ | 154903 |
| 1994 | $67^{2}$ | 557 | 15 | 1863 | $4^{2}$ | 141589 |  | $1^{2}$ | $1640{ }^{2}$ | $655{ }^{2}$ | 557 | 2 | 146950 |
| 1995 | $172{ }^{2}$ | 358 | 53 | 935 |  | 165001 |  | 5 | 1148 |  | 688 | 18 | 168378 |
| 1996 | $248{ }^{2}$ | 346 | 165 | 2615 |  | 166045 |  | 24 | 1159 | 6 | 707 | 33 | 171348 |
| 1997 | $193{ }^{2}$ | 560 | $363{ }^{2}$ | 2915 |  | 136927 |  | 12 | 1774 | 41 | 799 | 45 | 143629 |
| 1998 | 366 | 932 | $437{ }^{2}$ | 2936 |  | 144103 |  | 47 | 3836 | 275 | 355 | 40 | 153327 |
| 1999 | 181 | $638{ }^{2}$ | $655{ }^{2}$ | 2473 | 146 | 141941 |  | 17 | 3929 | 24 | 339 | 32 | 150375 |
| 2000 | $224{ }^{2}$ | 1438 | $651{ }^{2}$ | 2573 | 33 | 125932 |  | 46 | 4452 | 117 | 454 | $8^{2}$ | 135928 |
| 2001 | 537 | 1279 | $701{ }^{2}$ | 2690 | 57 | 124928 |  | 75 | 4951 | 119 | 514 | 2 | 135853 |
| 2002 | 788 | 1048 | 1393 | 2642 | 78 | 142941 |  | 118 | 5402 | 37 | 420 | 3 | 154870 |
| 2003 | 2056 | 1022 | $929{ }^{2}$ | 2763 | $80^{2}$ | 150400 |  | 147 | 3894 | 18 | 265 | $18^{2}$ | 161592 |
| 2004 | 3071 | 255 | $891{ }^{2}$ | 2161 | 319 | 147975 |  | 127 | 9192 | 87 | 544 | 14 | 164636 |
| 2005 | 3152 | 447 | $817{ }^{2}$ | 2048 | 395 | 162338 |  | 354 | 8362 | 25 | 630 |  | 178568 |
| 2006 | 1795 | 899 | $786{ }^{2}$ | 2779 | 255 | 195462 | 89 | $339{ }^{2}$ | 9823 | $21^{2}$ | 532 | 42 | 212822 |
| 2007 | 2048 | 966 | $810^{2}$ | 3019 | 219 | 178644 | 99 | 412 | 12168 | $53^{2}$ | 558 | 12 | 199008 |
| 2008 | 2314 | 1009 | $503{ }^{2}$ | 2263 | 113 | 165998 | 66 | 348 | 11577 | 33 | 506 | 10 | 184740 |
| $2009{ }^{1}$ | $1501^{2}$ | 323 | $697{ }^{2}$ | 2021 | $67^{2}$ | 144338 | 31 | 199 | 11895 | $2^{2}$ | 356 | 32 | 161462 |

1 Provisional figures.
2 As reported to Norwegian authorities.
3 USSR prior to 1991.
4 Includes Estonia.
5 Includes Denmark, Netherlands, Ireland and Sweden

Table 5.1.2 Saithe in Sub-areas I and II (Northeast Arctic).
Landings ('000 tonnes) by gear category.

| Year | Purse Seine | Trawl | Gill Net | Others | Total |
| :---: | :---: | ---: | ---: | ---: | :---: |
| 1977 | 75.2 | 69.5 | 19.3 | 12.7 | $176.7^{2}$ |
| 1978 | 62.9 | 57.6 | 21.1 | 13.9 | 155.5 |
| 1979 | 74.7 | 52.5 | 21.6 | 15.9 | 164.7 |
| 1980 | 61.3 | 46.8 | 21.1 | 15.4 | 144.6 |
| 1981 | 64.3 | 72.4 | 24.0 | 14.8 | 175.5 |
| 1982 | 76.4 | 59.4 | 16.7 | 15.5 | 168.0 |
| 1983 | 54.1 | 68.2 | 19.6 | 15.0 | 156.9 |
| 1984 | 36.4 | 85.6 | 23.7 | 13.1 | 158.8 |
| 1985 | 31.1 | 49.9 | 14.6 | 11.6 | 107.2 |
| 1986 | 7.9 | 36.2 | 12.3 | 8.2 | $64.6^{2}$ |
| 1987 | 34.9 | 27.7 | 19.0 | 10.8 | 92.4 |
| 1988 | 43.5 | 45.4 | 15.3 | 10.0 | 114.2 |
| 1989 | 49.5 | 45.0 | 16.9 | 11.4 | 122.8 |
| 1990 | 24.6 | 44.0 | 19.3 | 7.9 | 95.8 |
| 1991 | 38.9 | 40.1 | 18.9 | 9.4 | 107.3 |
| 1992 | 27.1 | 67.0 | 22.3 | 11.2 | 127.6 |
| 1993 | 33.1 | 84.9 | 21.2 | 15.7 | 154.9 |
| 1994 | 30.2 | 82.2 | 21.1 | 13.5 | $147.0{ }^{3}$ |
| 1995 | 21.8 | 103.5 | 26.9 | 16.1 | $168.4^{4}$ |
| 1996 | 46.9 | 72.5 | 31.6 | 20.3 | 171.3 |
| 1997 | 44.4 | 55.9 | 24.4 | 19.0 | 143.6 |
| 1998 | 44.4 | 57.7 | 27.6 | 23.6 | 153.3 |
| 1999 | 39.2 | 57.9 | 29.7 | 23.6 | 150.4 |
| 2000 | 28.3 | 54.5 | 29.6 | 23.5 | 135.9 |
| 2001 | 28.1 | 58.1 | 28.2 | 21.5 | 135.9 |
| 2002 | 27.4 | 75.5 | 30.4 | 21.5 | 154.8 |
| 2003 | 43.3 | 73.8 | 25.2 | 19.3 | 161.6 |
| 2004 | 41.8 | 74.6 | 26.9 | 21.3 | 164.6 |
| 2005 | 42.1 | 91.8 | 25.6 | 19.1 | 178.6 |
| 2006 | 73.5 | 87.1 | 29.7 | 22.5 | 212.8 |
| 2007 | 41.8 | 100.7 | 33.3 | 23.2 | 199.0 |
| 2008 | 39.4 | 91.2 | 37.0 | 17.1 | 184.7 |
| 2009 | 35.5 | 80.9 | 33.0 | 12.1 | 161.5 |
|  |  |  |  |  |  |

${ }^{1}$ Provisional figures.
${ }^{2}$ Unresolved discrepancy between Norwegian catch by gear figures and the total reported to ICES for these years.
${ }^{3}$ Includes 4,300 tonnes not categorized by gear, proportionally adjusted.
${ }^{4}$ Reduced by 1,200 tonnes not categorized by gear, proportionally adjusted.

Table 5.2.1 Saithe in Sub-areas I and II (Northeast Arctic).
Norwegian trawl CPUE by agegroup (Catch in numbers per trawlhour).
Shaded area shows indices applied in the assessment.

| Year |  | Agegroup |  |  |  |  |  |  |  | Total CPUE (kg/h) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | effort | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | Quarter 1-4 |  |
| 1994 | 1 | 3.4 | 83.2 | 280.2 | 174.0 | 24.0 | 5.3 | 1.7 | 3.3 | 575 |  |
| 1995 | 1 | 28.1 | 150.0 | 208.3 | 226.3 | 35.9 | 5.9 | 0.2 | 1.5 | 656 |  |
| 1996 | 1 | 17.0 | 84.7 | 113.2 | 164.7 | 217.1 | 24.9 | 5.3 | 0.5 | 628 |  |
| 1997 | 1 | 10.7 | 28.5 | 148.3 | 151.1 | 194.4 | 122.3 | 12.9 | 1.3 | 670 |  |
| 1998 | 1 | 2.4 | 24.5 | 41.1 | 181.6 | 69.2 | 42.1 | 12.1 | 5.7 | 379 |  |
| 1999 | 1 | 11.0 | 26.6 | 74.9 | 56.8 | 131.6 | 30.2 | 22.1 | 6.3 | 359 |  |
| 2000 | 1 | 5.4 | 58.8 | 62.9 | 117.9 | 91.3 | 122.6 | 46.4 | 52.4 | 558 |  |
| 2001 | 1 | 5.4 | 32.2 | 176.1 | 126.8 | 119.8 | 50.7 | 72.3 | 34.7 | 618 |  |
| 2002 | 1 | 6.9 | 52.2 | 84.9 | 264.3 | 59.6 | 61.2 | 28.0 | 52.1 | 609 |  |
| 2003 | 1 | 4.0 | 105.9 | 161.7 | 107.3 | 154.7 | 99.8 | 82.6 | 51.1 | 767 |  |
| 2004 | 1 | 2.4 | 5.8 | 141.8 | 105.4 | 135.3 | 169.6 | 54.5 | 74.8 | 690 |  |
| 2005 | 1 | 13.4 | 38.6 | 103.3 | 305.7 | 145.9 | 82.1 | 145.8 | 49.0 | 884 |  |
| 2006 | 1 | 0.3 | 53.5 | 99.2 | 86.9 | 202.3 | 116.9 | 103.9 | 97.7 | 761 |  |
| 2007 | 1 | 3.5 | 11.2 | 206.8 | 161.8 | 109.1 | 165.6 | 110.7 | 58.0 | 827 |  |
| 2008 | 1 | 15.8 | 81.1 | 46.3 | 266.0 | 149.1 | 90.8 | 135.6 | 83.9 | 868 |  |
| $2009{ }^{1}$ | 1 | 15.4 | 199.6 | 133.0 | 74.8 | 205.2 | 55.4 | 32.9 | 70.2 | 787 |  |

Table 5.2.2 Saithe in Sub-areas I and II (Northeast Arctic).
Acoustic abundance indices from Norwegian surveys in October-November. In 1985-1991 the area coverage was incomplete. Numbers in millions.
Shaded area shows indices applied in the assessment

| Year | Age |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6/6+ | 7 | 8 | 9 | 10+ | Total |
| 1985 | 3.1 | 4.9 | 2.4 | 0.5 | 0.0 |  |  |  |  | 10.9 |
| 1986 | 19.5 | 40.8 | 3.6 | 1.8 | 1.8 |  |  |  |  | 67.5 |
| 1987 | 1.8 | 22.0 | 48.4 | 1.8 | 1.7 |  |  |  |  | 75.7 |
| 1988 | 15.7 | 22.5 | 19.0 | 7.1 | 0.6 |  |  |  |  | 64.9 |
| 1989 | 24.8 | 28.4 | 17.0 | 10.1 | 12.4 |  |  |  |  | 92.7 |
| 1990 | 99.6 | 31.9 | 14.7 | 5.1 | 7.4 |  |  |  |  | 158.7 |
| 1991 | 87.8 | 104.0 | 4.6 | 4.0 | 7.1 |  |  |  |  | 207.5 |
| 1992 | 163.5 | 273.6 | 57.5 | 6.2 | 8.8 |  |  |  |  | 509.6 |
| 1993 | 106.9 | 227.7 | 103.9 | 12.7 | 3.2 |  |  |  |  | 454.4 |
| 1994 | 35.1 | 87.1 | 108.9 | 41.4 | 8.1 | 0.7 | 1.0 | 0.5 | 1.0 | 283.8 |
| 1995 | 38.4 | 166.1 | 86.5 | 46.5 | 16.5 | 2.4 | 0.0 | 0.0 | 1.0 | 357.5 |
| 1996 | 48.8 | 122.6 | 207.4 | 31.7 | 15.1 | 4.0 | 0.5 | 0.0 | 0.0 | 430.0 |
| 1997 | 5.5 | 38.0 | 184.8 | 79.8 | 50.6 | 9.6 | 1.2 | 0.0 | 0.3 | 369.8 |
| 1998 | 44.0 | 96.7 | 202.6 | 69.3 | 84.3 | 6.6 | 3.8 | 0.7 | 0.1 | 508.1 |
| 1999 | 61.1 | 233.8 | 72.9 | 62.2 | 21.0 | 19.2 | 5.9 | 1.4 | 0.4 | 477.8 |
| 2000 | 164.8 | 142.5 | 176.3 | 11.6 | 11.5 | 8.0 | 4.0 | 1.0 | 2.0 | 521.7 |
| 2001 | 104.7 | 275.9 | 45.9 | 53.8 | 5.6 | 6.1 | 3.2 | 3.4 | 1.9 | 500.5 |
| 2002 | 25.5 | 230.2 | 92.6 | 18.9 | 10.6 | 2.2 | 0.9 | 0.8 | 1.2 | 382.9 |
| 2003 | 31.0 | 87.5 | 151.7 | 26.1 | 6.2 | 6.4 | 1.2 | 0.7 | 1.3 | 312.1 |
| 2004 | 152.2 | 212.4 | 118.7 | 49.1 | 19.2 | 4.7 | 3.0 | 3.1 | 3.1 | 565.5 |
| 2005 | 22.2 | 228.1 | 67.2 | 20.3 | 16.5 | 7.7 | 2.2 | 1.7 | 0.9 | 366.7 |
| 2006 | 98.2 | 42.6 | 142.9 | 19.4 | 4.6 | 8.5 | 5.6 | 2.1 | 3.5 | 327.3 |
| 2007 | 45.4 | 111.0 | 27.1 | 61.1 | 7.9 | 5.8 | 4.1 | 4.3 | 1.1 | 267.9 |
| 2008 | 55.6 | 97.2 | 29.2 | 13.8 | 11.9 | 4.0 | 1.0 | 1.0 | 1.6 | 215.3 |
| 2009 | 52.9 | 139.8 | 80.2 | 7.7 | 5.2 | 6.8 | 0.9 | 0.7 | 1.7 | 295.9 |

Table 5.3.1 Northeast Arctic saithe. Norwegian sampling level in 2009 by ICES area and quarter.

|  | Ne A SAITHE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 |  | research vessels |  |  |  | commercial vessels |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  | Norwegian | Com mercial |
| ICES/NAFO |  | specimen samples |  | length samples |  | specimen samples |  | length samples |  | Landings in | samples |
| region | Q | samples | no | samples | no | samples | no | samples | no | Tonnes | $\begin{aligned} & \text { per } \\ & 1000 \end{aligned}$ |
| I | 1 | 4 | 17 | 15 | 216 | 1 | 29 | 12 | 71 | 187.2 | 69.5 |
|  | 2 | 2 | 20 | 65 | 830 | 3 | 124 | 47 | 1935 | 2340.5 | 21.4 |
|  | 3 | 9 | 63 | 80 | 2051 | 0 | 0 | 6 | 1189 | 6739.1 | 0.9 |
|  | 4 | 16 | 94 | 34 | 522 | 0 | 0 | 2 | 5 | 2278.3 | 0.9 |
|  | total | 31 | 194 | 194 | 3619 | 4 | 153 | 67 | 3200 | 11545.1 | 6.1 |
| IIa | 1 | 241 | 8858 | 455 | 15915 | 16 | 422 | 229 | 15642 | 55243.1 | 4.4 |
|  | 2 | 89 | 1836 | 216 | 5311 | 29 | 816 | 181 | 16288 | 27098.6 | 7.7 |
|  | 3 | 58 | 827 | 124 | 3827 | 9 | 142 | 192 | 11302 | 27646.5 | 7.3 |
|  | 4 | 98 | 1137 | 148 | 4778 | 5 | 106 | 59 | 5995 | 22736.8 | 2.8 |
|  | total | 486 | 12658 | 943 | 29831 | 59 | 1486 | 661 | 49227 | 132725.0 | 5.4 |
| IIb | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0.3 | 0 |
|  | 2 | 0 | 0 | 11 | 31 | 0 | 0 | 1 | 1 | 23.6 | 42.3 |
|  | 3 | 3 | 5 | 18 | 46 | 0 | 0 | 1 | 1 | 7.2 | 138.3 |
|  | 4 | 12 | 23 | 21 | 52 | 0 | 0 | 1 | 1 | 35.7 | 28.0 |
|  | total | 15 | 28 | 51 | 130 | 0 | 0 | 3 | 3 | 66.9 | 44.8 |

Table 5.3.2 Northeast Arctic saithe. Catch numbers at age Run title : North-East Arctic saithe At 22/04/2010 16:49

|  | Table 1 | Catch numbers at age |  |  |  | Numbers*10**-3 |  | 1966 | 1967 | 1968 | 1969 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAF | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 13517 | 25237 | 45932 | 51171 | 10925 | 42578 | 25127 | 28457 | 29955 | 76011 |
|  | 4 | 16828 | 12929 | 13720 | 35199 | 72344 | 5737 | 61199 | 23826 | 21856 | 11745 |
|  | 5 | 17422 | 17707 | 5449 | 7165 | 15966 | 30171 | 14727 | 34493 | 6065 | 16650 |
|  | 6 | 6514 | 5379 | 10218 | 5659 | 3299 | 11635 | 14475 | 3957 | 9846 | 4666 |
|  | 7 | 6281 | 1886 | 2991 | 4699 | 4214 | 3282 | 5220 | 5388 | 936 | 4716 |
|  | 8 | 3088 | 1371 | 1262 | 1337 | 3223 | 2421 | 1542 | 2797 | 2274 | 1107 |
|  | 9 | 1691 | 736 | 1156 | 1308 | 1518 | 3135 | 1047 | 1356 | 1070 | 1682 |
|  | 10 | 956 | 573 | 556 | 848 | 1482 | 802 | 1083 | 1340 | 686 | 663 |
|  | 11 | 481 | 538 | 611 | 550 | 1282 | 1136 | 530 | 814 | 465 | 199 |
|  | 12 | 363 | 275 | 369 | 467 | 965 | 652 | 628 | 603 | 284 | 138 |
|  | 13 | 260 | 112 | 282 | 399 | 561 | 509 | 670 | 528 | 168 | 30 |
|  | 14 | 185 | 89 | 224 | 166 | 443 | 802 | 497 | 391 | 156 | 47 |
|  | +gp | 673 | 726 | 643 | 580 | 1069 | 1023 | 929 | 1014 | 314 | 88 |
| 0 | TOTAL | 68259 | 67558 | 83413 | 109548 | 117291 | 103883 | 127674 | 104964 | 74075 | 117742 |
|  | TONSL | 133515 | 105951 | 120707 | 148627 | 197426 | 185600 | 203788 | 181326 | 110247 | 140060 |
|  | SOPCC | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
|  | Table 1 | Catch numbers at age |  |  |  | Numbers*10**-3 |  |  |  |  |  |
|  | YEAF | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 43834 | 61743 | 55351 | 62938 | 36884 | 70255 | 135592 | 105935 | 56505 | 75819 |
|  | 4 | 63270 | 47522 | 44490 | 20793 | 44149 | 13502 | 33159 | 36703 | 31946 | 28545 |
|  | 5 | 14081 | 21614 | 24752 | 22199 | 15714 | 18901 | 8618 | 10845 | 14396 | 17280 |
|  | 6 | 16298 | 7661 | 8650 | 13224 | 20476 | 5123 | 9448 | 2205 | 5232 | 5384 |
|  | 7 | 5157 | 7690 | 4769 | 5868 | 12182 | 9018 | 3725 | 4633 | 1694 | 3550 |
|  | 8 | 8004 | 2326 | 3012 | 3246 | 4815 | 7841 | 3483 | 1557 | 2132 | 1178 |
|  | 9 | 2521 | 3489 | 1584 | 2368 | 3267 | 3365 | 2905 | 1718 | 1082 | 1659 |
|  | 10 | 3722 | 1760 | 1817 | 2153 | 2512 | 2714 | 1870 | 1030 | 1126 | 536 |
|  | 11 | 1103 | 2514 | 1044 | 1291 | 1440 | 2237 | 1183 | 495 | 756 | 373 |
|  | 12 | 762 | 1045 | 676 | 653 | 1448 | 1438 | 924 | 261 | 786 | 344 |
|  | 13 | 325 | 284 | 281 | 670 | 433 | 530 | 530 | 226 | 328 | 206 |
|  | 14 | 278 | 186 | 222 | 365 | 264 | 300 | 152 | 62 | 267 | 272 |
|  | +gp | 349 | 373 | 452 | 259 | 247 | 276 | 334 | 169 | 345 | 264 |
| 0 | TOTAL | 159704 | 158207 | 147100 | 136027 | 143831 | 135500 | 201923 | 165839 | 116595 | 135410 |
|  | TONSL | 264924 | 241272 | 214334 | 213859 | 264121 | 233453 | 242486 | 182817 | 155464 | 164680 |
|  | SOPCC | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
|  | Table 1 | Catch numbers at age |  |  |  | Numbers*10**-3 |  |  |  |  |  |
|  | YEAF | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 40303 | 85966 | 35853 | 18216 | 43579 | 48989 | 21322 | 18555 | 8144 | 12607 |
|  | 4 | 36202 | 22345 | 67150 | 25108 | 34927 | 11992 | 12433 | 51742 | 35928 | 19400 |
|  | 5 | 9100 | 22044 | 13481 | 34543 | 12679 | 7200 | 5845 | 4506 | 32901 | 33343 |
|  | 6 | 6302 | 3706 | 8477 | 3408 | 11775 | 5287 | 4363 | 3238 | 4570 | 18578 |
|  | 7 | 3161 | 2611 | 1088 | 3178 | 1193 | 3746 | 2704 | 3624 | 2333 | 1762 |
|  | 8 | 1322 | 2056 | 1291 | 1243 | 1862 | 776 | 1349 | 784 | 1222 | 352 |
|  | 9 | 145 | 378 | 476 | 803 | 589 | 879 | 338 | 644 | 968 | 177 |
|  | 10 | 721 | 286 | 271 | 261 | 585 | 134 | 438 | 267 | 321 | 189 |
|  | 11 | 406 | 258 | 124 | 215 | 407 | 274 | 123 | 263 | 73 | 1 |
|  | 12 | 449 | 91 | 116 | 130 | 158 | 214 | 65 | 164 | 12 | 149 |
|  | 13 | 254 | 147 | 78 | 170 | 123 | 55 | 30 | 154 | 2 | 0 |
|  | 14 | 236 | 97 | 100 | 99 | 179 | 126 | 54 | 102 | 15 | 36 |
|  | +gp | 265 | 50 | 44 | 188 | 77 | 32 | 3 | 145 | 1 | 20 |
| 0 | TOTAL | 98866 | 140035 | 128549 | 87562 | 108133 | 79704 | 49067 | 84188 | 86490 | 86614 |
|  | TONSL | 144554 | 175540 | 168034 | 156936 | 158786 | 107183 | 67396 | 92391 | 114242 | 122817 |
|  | SOPCC | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 105 |

Table 5.3.2 continue

|  | Table 1 | Catch numbers at age |  |  | Numbers*10** ${ }^{\text {a }}$ |  |  |  | 1997 | 1998 | 1999 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAF | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 23792 | 68682 | 44627 | 22812 | 7063 | 17178 | 10510 | 11789 | 3091 | 9655 |
|  | 4 | 16930 | 13630 | 33294 | 61931 | 32671 | 52109 | 54886 | 11698 | 16215 | 12236 |
|  | 5 | 9054 | 5752 | 5987 | 31102 | 49410 | 40145 | 18499 | 35011 | 11946 | 22872 |
|  | 6 | 10238 | 4883 | 5412 | 3747 | 19058 | 30451 | 18357 | 13567 | 31818 | 10347 |
|  | 7 | 7341 | 3877 | 4751 | 1759 | 2058 | 4177 | 17834 | 13452 | 8376 | 18930 |
|  | 8 | 1076 | 2381 | 3176 | 1378 | 724 | 483 | 2849 | 7058 | 5539 | 3374 |
|  | 9 | 160 | 383 | 1462 | 1027 | 421 | 125 | 485 | 812 | 2873 | 3343 |
|  | 10 | 112 | 61 | 286 | 797 | 278 | 259 | 214 | 55 | 727 | 2290 |
|  | 11 | 150 | 90 | 93 | 76 | 528 | 31 | 148 | 48 | 111 | 419 |
|  | 12 | 37 | 68 | 46 | 35 | 92 | 176 | 68 | 42 | 65 | 103 |
|  | 13 | 31 | 1 | 163 | 1 | 13 | 2 | 196 | 27 | 19 | 24 |
|  | 14 | 0 | 12 | 0 | 17 | 15 | 42 | 59 | 21 | 0 | 11 |
|  | +gp | 50 | 8 | 141 | 18 | 9 | 43 | 2 | 8 | 198 | 32 |
| 0 | TOTAL | 68971 | 99828 | 99438 | 124700 | 112340 | 145221 | 124107 | 93588 | 80978 | 83636 |
|  | TONSL | 95848 | 107327 | 127604 | 154903 | 146950 | 168378 | 171348 | 143629 | 153327 | 150375 |
|  | SOPCC | 102 | 101 | 105 | 101 | 100 | 100 | 100 | 100 | 100 | 100 |
|  | Table 1 | Catch numbers at age |  |  |  | Numbers*10**-3 |  |  |  |  |  |
|  | YEAF | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 9175 | 3816 | 6582 | 2345 | 1002 | 26093 | 1590 | 3144 | 25259 | 5698 |
|  | 4 | 22768 | 7946 | 17492 | 50653 | 6129 | 12543 | 68137 | 4115 | 18953 | 38011 |
|  | 5 | 7747 | 26960 | 11573 | 13600 | 33840 | 9841 | 12328 | 39889 | 5969 | 9704 |
|  | 6 | 10676 | 8769 | 25671 | 7123 | 10613 | 23141 | 10098 | 15301 | 24363 | 6411 |
|  | 7 | 6123 | 7120 | 5312 | 9594 | 7494 | 10799 | 16757 | 7963 | 9712 | 16220 |
|  | 8 | 8303 | 3146 | 4276 | 5494 | 8307 | 5659 | 8080 | 11302 | 5624 | 4813 |
|  | 9 | 2530 | 4687 | 2382 | 3545 | 2792 | 7852 | 5671 | 7749 | 7697 | 2982 |
|  | 10 | 2652 | 1935 | 3431 | 2519 | 3088 | 2674 | 5127 | 4138 | 4705 | 3991 |
|  | 11 | 1022 | 1406 | 965 | 2327 | 2377 | 713 | 1815 | 2157 | 1606 | 2343 |
|  | 12 | 151 | 433 | 1016 | 1112 | 2057 | 387 | 1013 | 505 | 1163 | 1006 |
|  | 13 | 8 | 60 | 281 | 420 | 338 | 465 | 733 | 254 | 145 | 236 |
|  | 14 | 25 | 8 | 68 | 170 | 536 | 357 | 506 | 52 | 108 | 93 |
|  | +gp | 13 | 27 | 55 | 111 | 141 | 379 | 277 | 38 | 156 | 103 |
| 0 | TOTAL | 71193 | 66313 | 79104 | 99013 | 78714 | 100903 | 132132 | 96607 | 105460 | 91611 |
|  | TONSL | 135928 | 135853 | 154870 | 161592 | 164636 | 178568 | 212822 | 199008 | 184740 | 161462 |
|  | SOPCC | 101 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 102 | 100 |


| Table 5.3.3 Northeast Arctic saithe. Catch weight at age |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Table 2 Catch weights at age (kg) |  |  |  |  |  |  |  |  |  |  |  |
|  | YEAF | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0.71 | 0.71 | 0.71 | 0.71 | 0.71 | 0.71 | 0.71 | 0.71 | 0.71 | 0.71 |
|  | 4 | 1.11 | 1.11 | 1.11 | 1.11 | 1.11 | 1.11 | 1.11 | 1.11 | 1.11 | 1.11 |
|  | 5 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 |
|  | 6 | 2.33 | 2.33 | 2.33 | 2.33 | 2.33 | 2.33 | 2.33 | 2.33 | 2.33 | 2.33 |
|  | 7 | 3.16 | 3.16 | 3.16 | 3.16 | 3.16 | 3.16 | 3.16 | 3.16 | 3.16 | 3.16 |
|  | 8 | 4.03 | 4.03 | 4.03 | 4.03 | 4.03 | 4.03 | 4.03 | 4.03 | 4.03 | 4.03 |
|  | 9 | 4.87 | 4.87 | 4.87 | 4.87 | 4.87 | 4.87 | 4.87 | 4.87 | 4.87 | 4.87 |
|  | 10 | 5.63 | 5.63 | 5.63 | 5.63 | 5.63 | 5.63 | 5.63 | 5.63 | 5.63 | 5.63 |
|  | 11 | 6.44 | 6.44 | 6.44 | 6.44 | 6.44 | 6.44 | 6.44 | 6.44 | 6.44 | 6.44 |
|  | 12 | 7.11 | 7.11 | 7.11 | 7.11 | 7.11 | 7.11 | 7.11 | 7.11 | 7.11 | 7.11 |
|  | 13 | 7.82 | 7.82 | 7.82 | 7.82 | 7.82 | 7.82 | 7.82 | 7.82 | 7.82 | 7.82 |
|  | 14 | 8.92 | 8.92 | 8.92 | 8.92 | 8.92 | 8.92 | 8.92 | 8.92 | 8.92 | 8.92 |
|  | +gp | 9.5 | 9.5 | 9.5 | 9.5 | 9.5 | 9.5 | 9.5 | 9.5 | 9.5 | 9.5 |
| 0 | SOPC | 1 | 1 | 1.0001 | 1 | 1 | 1 | 1 | 1 | 0.9999 | 1 |
| Table 2 Catch weights at age (kg) |  |  |  |  |  |  |  |  |  |  |  |
|  | YEAF | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0.71 | 0.71 | 0.71 | 0.71 | 0.71 | 0.71 | 0.71 | 0.71 | 0.71 | 0.71 |
|  | 4 | 1.11 | 1.11 | 1.11 | 1.11 | 1.11 | 1.11 | 1.11 | 1.11 | 1.11 | 1.11 |
|  | 5 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 |
|  | 6 | 2.33 | 2.33 | 2.33 | 2.33 | 2.33 | 2.33 | 2.33 | 2.33 | 2.33 | 2.33 |
|  | 7 | 3.16 | 3.16 | 3.16 | 3.16 | 3.16 | 3.16 | 3.16 | 3.16 | 3.16 | 3.16 |
|  | 8 | 4.03 | 4.03 | 4.03 | 4.03 | 4.03 | 4.03 | 4.03 | 4.03 | 4.03 | 4.03 |
|  | 9 | 4.87 | 4.87 | 4.87 | 4.87 | 4.87 | 4.87 | 4.87 | 4.87 | 4.87 | 4.87 |
|  | 10 | 5.63 | 5.63 | 5.63 | 5.63 | 5.63 | 5.63 | 5.63 | 5.63 | 5.63 | 5.63 |
|  | 11 | 6.44 | 6.44 | 6.44 | 6.44 | 6.44 | 6.44 | 6.44 | 6.44 | 6.44 | 6.44 |
|  | 12 | 7.11 | 7.11 | 7.11 | 7.11 | 7.11 | 7.11 | 7.11 | 7.11 | 7.11 | 7.11 |
|  | 13 | 7.82 | 7.82 | 7.82 | 7.82 | 7.82 | 7.82 | 7.82 | 7.82 | 7.82 | 7.82 |
|  | 14 | 8.92 | 8.92 | 8.92 | 8.92 | 8.92 | 8.92 | 8.92 | 8.92 | 8.92 | 8.92 |
|  | +gp | 9.5 | 9.5 | 9.5 | 9.5 | 9.5 | 9.5 | 9.5 | 9.5 | 9.5 | 9.5 |
| 0 | SOPC | 1 | 0.9999 | 1 | 0.9996 | 1 | 1 | 1 | 1 | 1 | 1 |
| Table 2 Catch weights at age (kg) |  |  |  |  |  |  |  |  |  |  |  |
|  | YEAF | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0.79 | 0.73 | 0.77 | 1.05 | 0.71 | 0.75 | 0.59 | 0.53 | 0.62 | 0.74 |
|  | 4 | 1.27 | 1.4 | 1.12 | 1.33 | 1.26 | 1.33 | 1.22 | 0.84 | 0.87 | 0.95 |
|  | 5 | 2.03 | 2.05 | 2.02 | 1.86 | 2.02 | 2.07 | 1.97 | 1.66 | 1.31 | 1.4 |
|  | 6 | 2.55 | 2.76 | 2.61 | 2.8 | 2.7 | 2.63 | 2.3 | 2.32 | 2.43 | 1.78 |
|  | 7 | 3.29 | 3.3 | 3.27 | 4 | 3.88 | 3.28 | 2.87 | 2.97 | 3.87 | 2.96 |
|  | 8 | 4.34 | 4.38 | 3.91 | 4.18 | 4.47 | 3.96 | 3.72 | 4 | 5.38 | 3.73 |
|  | 9 | 5.15 | 5.95 | 4.69 | 5.33 | 5.36 | 4.54 | 4.3 | 4.72 | 5.83 | 4.62 |
|  | 10 | 5.75 | 6.39 | 5.63 | 5.68 | 6.06 | 5.55 | 4.69 | 5.44 | 5.36 | 4.66 |
|  | 11 | 6.11 | 6.61 | 7.18 | 7.31 | 6.28 | 6.88 | 5.84 | 5.79 | 6.92 | 8.34 |
|  | 12 | 5.94 | 6.88 | 7.21 | 8.68 | 6.89 | 8.14 | 6.39 | 6.28 | 8.72 | 6.77 |
|  | 13 | 6.64 | 6.75 | 7 | 8.54 | 8.2 | 6.06 | 8.11 | 7.02 | 7.88 | 10.04 |
|  | 14 | 7.73 | 7.13 | 8.03 | 8.57 | 9.14 | 9.66 | 7.55 | 8.36 | 8.94 | 9.13 |
|  | +gp | 9.47 | 7.66 | 9.44 | 10.37 | 6.47 | 13.72 | 10.08 | 8.48 | 10 | 11.95 |
| 0 | SOPC | 1 | 0.9999 | 1 | 1 | 0.9999 | 0.9997 | 1 | 0.9999 | 0.9999 | 1.0469 |

Table 5.3.3 continue

|  | Table 2 | Catch | ights at | (kg) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAF | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0.71 | 0.68 | 0.67 | 0.61 | 0.52 | 0.56 | 0.59 | 0.62 | 0.68 | 0.67 |
|  | 4 | 1 | 1.05 | 1.01 | 0.99 | 0.76 | 0.79 | 0.82 | 0.95 | 1 | 1.05 |
|  | 5 | 1.45 | 1.85 | 1.92 | 1.65 | 1.24 | 1.19 | 1.33 | 1.24 | 1.48 | 1.45 |
|  | 6 | 2.09 | 2.39 | 2.28 | 2.46 | 2.12 | 1.71 | 1.84 | 1.72 | 1.87 | 1.93 |
|  | 7 | 2.49 | 3.08 | 2.77 | 2.85 | 3.22 | 2.87 | 2.48 | 2.35 | 2.58 | 2.27 |
|  | 8 | 3.75 | 3.35 | 3.2 | 3.03 | 3.83 | 3.78 | 3.73 | 3.1 | 3.07 | 2.97 |
|  | 9 | 3.9 | 4.48 | 3.73 | 3.71 | 4.69 | 4.06 | 4.32 | 4.19 | 4.13 | 3.61 |
|  | 10 | 6.74 | 4.66 | 6.35 | 4.49 | 5.31 | 5.3 | 5.34 | 5.79 | 5.44 | 4.1 |
|  | 11 | 4.94 | 5.62 | 6.9 | 5.56 | 5.66 | 6.86 | 5.98 | 6.77 | 6.7 | 4.93 |
|  | 12 | 4.93 | 6.3 | 7.18 | 6.56 | 6.91 | 6.59 | 6.26 | 6.62 | 4.97 | 6.59 |
|  | 13 | 8.2 | 6.73 | 6.88 | 10.56 | 6.3 | 7.88 | 7.36 | 7.3 | 5.23 | 7.52 |
|  | 14 | 8.2 | 11.55 | 7.5 | 6.73 | 9.45 | 9.16 | 9.61 | 9.15 | 6.8 | 7.88 |
|  | +gp | 8.59 | 9.58 | 9.14 | 8.41 | 8.95 | 10.53 | 13.64 | 11.48 | 10.1 | 7.46 |
| 0 | SOPC | 1.0235 | 1.0087 | 1.0517 | 1.0107 | 1 | 0.999 | 1.0019 | 1.0011 | 1.0015 | 1.0015 |
| Table 2 Catch weights at age (kg) |  |  |  |  |  |  |  |  |  |  |  |
|  | YEAF | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0.6 | 0.75 | 0.69 | 0.66 | 0.7 | 0.59 | 0.63 | 0.73 | 0.63 | 0.67 |
|  | 4 | 1.03 | 1.12 | 1.01 | 0.91 | 1.03 | 0.89 | 0.83 | 1.08 | 0.98 | 1.01 |
|  | 5 | 1.63 | 1.54 | 1.5 | 1.42 | 1.37 | 1.49 | 1.43 | 1.41 | 1.38 | 1.69 |
|  | 6 | 2.1 | 2.04 | 1.97 | 1.89 | 1.9 | 2.09 | 1.78 | 1.86 | 1.92 | 2.01 |
|  | 7 | 2.67 | 2.6 | 2.54 | 2.54 | 2.41 | 2.16 | 2.27 | 2.43 | 2.31 | 2.38 |
|  | 8 | 3.14 | 3.14 | 3.25 | 2.58 | 2.98 | 2.99 | 2.73 | 2.94 | 2.83 | 2.68 |
|  | 9 | 3.81 | 3.63 | 3.77 | 3.49 | 3.44 | 3.24 | 3.02 | 3.35 | 3.16 | 3.23 |
|  | 10 | 4.41 | 4.54 | 4.31 | 3.75 | 3.73 | 3.82 | 3.9 | 3.66 | 3.43 | 3.43 |
|  | 11 | 5.76 | 5.05 | 4.91 | 4.12 | 4.14 | 3.92 | 4.06 | 4.17 | 3.82 | 3.47 |
|  | 12 | 7.3 | 5.82 | 5.69 | 5.27 | 5.09 | 5.14 | 5.05 | 5.04 | 4.09 | 4.23 |
|  | 13 | 9.95 | 6.4 | 6.19 | 5.94 | 5.96 | 6.26 | 5.79 | 6.07 | 5.03 | 4.87 |
|  | 14 | 10.56 | 7.88 | 7.56 | 6.49 | 5.99 | 6.76 | 6.01 | 5.23 | 5.97 | 5.59 |
|  | +gp | 11.08 | 10.84 | 11.71 | 11.21 | 7.91 | 6.62 | 8.35 | 9.14 | 8.56 | 7.31 |
| 0 | SOPC | 1.0051 | 1.001 | 1.001 | 1.0033 | 1.0031 | 1.0026 | 1.0017 | 1.0009 | 1.0155 | 1.0021 |

Table 5.3.4. Saithe in Subareas I and II (Northeast Arctic). 3-year running average maturity ogive 1985-2009.

| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | $15+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1985 | 0.00 | 0.02 | 0.50 | 0.92 | 0.99 | 1.00 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1986 | 0.00 | 0.02 | 0.51 | 0.94 | 0.99 | 1.00 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1987 | 0.00 | 0.00 | 0.35 | 0.98 | 1.00 | 1.00 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1988 | 0.00 | 0.00 | 0.25 | 0.96 | 1.00 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1989 | 0.00 | 0.00 | 0.15 | 0.92 | 1.00 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1990 | 0.00 | 0.00 | 0.20 | 0.85 | 0.99 | 1.00 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1991 | 0.00 | 0.02 | 0.25 | 0.84 | 0.98 | 1.00 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1992 | 0.00 | 0.02 | 0.30 | 0.83 | 0.93 | 0.92 | 0.90 | 0.95 | 1 | 1 | 1 | 1 | 1 |
| 1993 | 0.00 | 0.02 | 0.26 | 0.88 | 0.92 | 0.89 | 0.87 | 0.89 | 1 | 0.98 | 1 | 1 | 1 |
| 1994 | 0.00 | 0.02 | 0.26 | 0.84 | 0.90 | 0.82 | 0.87 | 0.89 | 1 | 0.98 | 1 | 1 | 1 |
| 1995 | 0.00 | 0.02 | 0.22 | 0.80 | 0.92 | 0.90 | 0.97 | 0.94 | 1 | 0.98 | 1 | 1 | 1 |
| 1996 | 0.00 | 0.03 | 0.21 | 0.65 | 0.91 | 0.93 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1997 | 0.00 | 0.03 | 0.14 | 0.45 | 0.83 | 0.94 | 0.93 | 0.97 | 1 | 1 | 1 | 1 | 1 |
| 1998 | 0.00 | 0.04 | 0.07 | 0.33 | 0.74 | 0.93 | 0.92 | 0.96 | 1 | 1 | 1 | 1 | 1 |
| 1999 | 0.00 | 0.00 | 0.08 | 0.32 | 0.74 | 0.92 | 0.92 | 0.96 | 0.99 | 0.97 | 1 | 1 | 1 |
| 2000 | 0.00 | 0.00 | 0.08 | 0.46 | 0.82 | 0.96 | 0.98 | 0.99 | 0.97 | 0.94 | 1 | 1 | 1 |
| 2001 | 0.00 | 0.00 | 0.11 | 0.64 | 0.93 | 0.97 | 0.98 | 0.99 | 0.97 | 0.93 | 1 | 1 | 1 |
| 2002 | 0.00 | 0.00 | 0.13 | 0.78 | 0.95 | 0.98 | 0.98 | 0.99 | 0.98 | 0.96 | 1 | 1 | 1 |
| 2003 | 0.00 | 0.00 | 0.14 | 0.82 | 0.96 | 0.98 | 0.98 | 0.99 | 1.00 | 0.98 | 1 | 1 | 1 |
| 2004 | 0.00 | 0.00 | 0.21 | 0.80 | 0.97 | 0.99 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2005 | 0.00 | 0.03 | 0.30 | 0.82 | 0.97 | 0.99 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2006 | 0.00 | 0.04 | 0.40 | 0.86 | 0.98 | 0.99 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2007 | 0.00 | 0.05 | 0.42 | 0.87 | 0.97 | 0.98 | 1 | 0.97 | 1 | 1 | 1 | 1 | 1 |
| 2008 | 0.00 | 0.06 | 0.33 | 0.84 | 0.96 | 0.99 | 1 | 0.97 | 1 | 1 | 1 | 1 | 1 |
| 2009 | 0.00 | 0.07 | 0.30 | 0.81 | 0.95 | 0.98 | 1 | 0.95 | 1 | 1 | 1 | 1 | 1 |
|  |  |  |  |  |  |  |  |  |  |  |  | 1 |  |

Table 5.3.5 Northeast Arctic saithe. Tuning data sets applied in final XSA run

```
Northeast Arctic saithe (Sub-areas I and II)
104
FLT11: Nor trawl revised 2010 (Catch: Unknown) (Effort: Un-
known)
1994 2001
1 1 0.00 1.00
4
    1 83.2 280.2 174.0 24.0 5.3
    1 150.0 208.3 226.3 35.9 5.9
    1 84.7 113.2 164.7 217.1 24.9
    1 28.5 148.3 151.1 194.4 122.3
    1 24.5 41.1 181.6 69.2 42.1
    1 26.6 74.9 56.8 131.6 30.2
    1 5%.8 62.9 117.9 91.3 122.6
    1 32.2 176.1 126.8 119.8 50.7
FLT12: Nor trawl revised 2010 (Catch: Unknown) (Effort: Un-
known)
2002 2009
1 1 0.00 1.00
4
    1 52.2 
    1 105.9 161.7 107.3 154.7 99.8
    1 5.8 141.8 105.4 135.3 169.6
    1 38.6 103.3 305.7 145.9 82.1
    1 53.5 99.2 86.9 202.3 116.9
    1 11.2 206.8 161.8 109.1 165.6
    1 81.1 46.3 266.0 149.1 
    1
FLT13: Norway Ac Survey (Catch: Unknown) (Effort: Unknown)
1994 2001
1 1 0.75 0.85
3}
    1 87.1 108.9 41.4 8.1 0.1 0.7
    1 1rrrrre
    1 122.6 207.4 31.7 15.1 4.0
    1 38.0 184.8 79.8 50.6 90.6
    1
    1 233.8 72.9 62.2 21.0 19.2
    1 142.5 176.3 11.6 11.5 8.0
    1 275.9 45.9 53.8 5.0.6 6.1
FLT14: Norway Ac Survey (Catch: Unknown) (Effort: Unknown)
2002 2009
1 1 0.75 0.85
3
    1 230.2 92.6 18.9 10.6 2. 2 .2
    1 87.5 151.7 26.1 
    1 212.4 118.7 49.1 19.2 4.7
    1 228.1 
    1 42.6 142.9 19.4 4.6 4.6 8.5
    1 111.0 
    1 
    rrrrrr
```

Table 5.4.1. Northeast Arctic saithe. Data and parameter settings of exploratory and final XSAruns. Changes compared to 2009-assessment in bold.

| Run No. | 1 | 2 | 3 | 4 |
| :---: | :---: | :---: | :---: | :---: |
| Ass. type | SPALY | SFT | SFT | FINAL |
| Catch data | 1960-08 | 1960-09 | 1960-09 | 1960-09 |
| Age range | 3-11+ | 3-15+ | 3-15+ | 3-15+ |
| F bar | 4-7 | 4-7 | 4-7 | 4-7 |
| Fleet 11 Norw. trawl |  | $\begin{aligned} & \text { 1994-2001 } \\ & \text { age 4-8 } \\ & \text { Q1-4 } \end{aligned}$ |  | $\begin{aligned} & \text { 1994-2001 } \\ & \text { age 4-8 } \\ & \text { Q1-4 } \end{aligned}$ |
| Fleet 12 Norw. trawl | $\begin{aligned} & \hline \text { 1994-2006 } \\ & \text { age 4-8 } \\ & \text { Q2-4 } \\ & \hline \end{aligned}$ | 2002-2009 <br> age 4-8 <br> Q1-4 |  | 2002-2009 <br> age 4-8 <br> Q1-4 |
| Fleet 13 ac. survey | $\begin{aligned} & 1994-07 \\ & \text { age 3-7 } \end{aligned}$ |  | 1994-2001 <br> age 3-7 | 1994-2001 <br> age 3-7 |
| Fleet 14 ac. survey |  |  | 2002-2009 <br> age 3-7 | 2002-2009 <br> age 3-7 |
| Time series weights | Tricubic over 20y | No | No | No |
| Power model | No | No | No | No |
| Catchability (q) plateau | 8 | 8 | 8 | 8 |
| Survivorest. shrunk tow. Mean of | 5 years <br> 5 oldest ages | 5 years <br> 5 oldest ages | 5 years <br> 5 oldest ages | 5 years <br> 5 oldest ages |
| SE of mean | 0.5 | 1.5 | 1.5 | 1.5 |
| Min. fleetSE for pop. Est. | 0.3 | 0.3 | 0.3 | 0.3 |
| Prior weight. | None | None | None | None |

Table 5.5.1. Northeast Arctic saithe. Tuning diagnostics
Lowestoft VPA Version 3.1
22/04/2010 16:47
Extended Survivors Analysis
North-East Arctic saithe
CPUE data from file flt-split.dat
Catch data for 50 years. 1960 to 2009. Ages 3 to 15 .

| Fleet |  | Last year | First age |  | Last age |  | Alpha | Beta |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FLT11: Nc | 1994 | 2009 |  | 4 |  | 8 | 0 | 1 |
| FLT12: Nc | 2002 | 2009 |  | 4 |  | 8 | 0 | 1 |
| FLT13: Nc | 1994 | 2009 |  | 3 |  | 7 | 0.75 | 0.85 |
| FLT14: Nc | 2002 | 2009 |  | 3 |  | 7 | 0.75 | 0.85 |

Time series weights :
Tapered time weighting not applied

Catchability analysis :
Catchability independent of stock size for all ages
Catchability independent of age for ages >= 8

Terminal population estimation :
Survivor estimates shrunk towards the mean F of the final 5 years or the 5 oldest ages.
S.E. of the mean to which the estimates are shrunk $=1.500$

Minimum standard error for population estimates derived from each fleet $=.300$

Prior weighting not applied

Tuning had not converged after 30 iterations

Total absolute residual between iterations
29 and $30=.01747$

| Final year $F$ values |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| Iteration 2 | 0.0335 | 0.3218 | 0.2156 | 0.2449 | 0.279 | 0.2542 | 0.2397 | 0.3023 | 0.617 | 1.1183 |
| Iteration 3 | 0.0334 | 0.3216 | 0.2154 | 0.2446 | 0.2787 | 0.2539 | 0.2393 | 0.3013 | 0.6113 | 1.1151 |


| Age | 13 | 14 |
| :--- | ---: | ---: |
| Iteration 2 | 0.63 | 0.6989 |
| Iteration 3 | 0.6275 | 0.6956 |

1

Regression weights

Table 5.5.1. Continued

| Fishing mortalities <br> Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |  |
| 3 | 0.071 | 0.021 | 0.021 | 0.019 | 0.007 | 0.075 | 0.023 | 0.032 | 0.139 | 0.033 |
| 4 | 0.131 | 0.081 | 0.129 | 0.222 | 0.064 | 0.115 | 0.285 | 0.077 | 0.27 | 0.322 |
| 5 | 0.139 | 0.225 | 0.163 | 0.14 | 0.227 | 0.14 | 0.157 | 0.269 | 0.153 | 0.215 |
| 6 | 0.261 | 0.23 | 0.348 | 0.143 | 0.154 | 0.24 | 0.208 | 0.299 | 0.262 | 0.245 |
| 7 | 0.254 | 0.278 | 0.213 | 0.211 | 0.219 | 0.232 | 0.274 | 0.252 | 0.315 | 0.279 |
| 8 | 0.301 | 0.2 | 0.268 | 0.356 | 0.286 | 0.256 | 0.272 | 0.3 | 0.284 | 0.254 |
| 9 | 0.253 | 0.277 | 0.229 | 0.373 | 0.309 | 0.481 | 0.443 | 0.457 | 0.344 | 0.239 |
| 10 | 0.345 | 0.313 | 0.336 | 0.404 | 0.656 | 0.55 | 0.677 | 0.686 | 0.561 | 0.301 |
| 11 | 0.379 | 0.31 | 0.254 | 0.401 | 0.854 | 0.303 | 0.936 | 0.689 | 0.63 | 0.611 |
| 12 | 0.265 | 0.272 | 0.386 | 0.522 | 0.76 | 0.312 | 0.953 | 0.748 | 1.058 | 1.115 |
| 13 | 0.139 | 0.159 | 0.285 | 0.272 | 0.294 | 0.377 | 1.888 | 0.67 | 0.495 | 0.628 |
| 14 | 0.273 | 0.201 | 0.273 | 0.279 | 0.668 | 0.58 | 0.94 | 0.665 | 0.684 | 0.696 |

1
XSA population numbers (Thousands)

|  | AGE | AGEAR | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

$2000 \quad 1.48 \mathrm{E}+05 \quad 2.05 \mathrm{E}+05 \quad 6.61 \mathrm{E}+04 \quad 5.14 \mathrm{E}+04 \quad 3.02 \mathrm{E}+04 \quad 3.53 \mathrm{E}+04 \quad 1.25 \mathrm{E}+04 \quad 1.01 \mathrm{E}+04 \quad 3.58 \mathrm{E}+03 \quad 7.17 \mathrm{E}+02$ $20012.00 \mathrm{E}+05 \quad 1.13 \mathrm{E}+05 \quad 1.48 \mathrm{E}+05 \quad 4.71 \mathrm{E}+04 \quad 3.24 \mathrm{E}+04 \quad 1.92 \mathrm{E}+04 \quad 2.14 \mathrm{E}+04 \quad 7.95 \mathrm{E}+03 \quad 5.83 \mathrm{E}+03 \quad 2.01 \mathrm{E}+03$ 2002 3.50E+05 1.60E+05 $8.51 \mathrm{E}+049.65 \mathrm{E}+043.06 \mathrm{E}+04 \quad 2.01 \mathrm{E}+04 \quad 1.29 \mathrm{E}+041.33 \mathrm{E}+04 \quad 4.76 \mathrm{E}+03 \quad 3.50 \mathrm{E}+03$ $20031.35 \mathrm{E}+05 \quad 2.81 \mathrm{E}+05 \quad 1.15 \mathrm{E}+05 \quad 5.92 \mathrm{E}+04 \quad 5.57 \mathrm{E}+04 \quad 2.03 \mathrm{E}+04 \quad 1.26 \mathrm{E}+04 \quad 8.37 \mathrm{E}+03 \quad 7.79 \mathrm{E}+03 \quad 3.02 \mathrm{E}+03$ 2004 1.58E+05 1.09E+05 1.84E+05 8.22E+04 4.21E+04 3.70E+04 1.16E+04 7.09E+03 4.57E+03 4.27E+03 $2005 \quad 4.00 \mathrm{E}+05 \quad 1.28 \mathrm{E}+05 \quad 8.34 \mathrm{E}+04 \quad 1.20 \mathrm{E}+05 \quad 5.77 \mathrm{E}+04 \quad 2.76 \mathrm{E}+04 \quad 2.27 \mathrm{E}+04 \quad 6.99 \mathrm{E}+03 \quad 3.01 \mathrm{E}+03 \quad 1.59 \mathrm{E}+03$ $20067.66 \mathrm{E}+04 \quad 3.04 \mathrm{E}+05 \quad 9.35 \mathrm{E}+04 \quad 5.94 \mathrm{E}+04 \quad 7.74 \mathrm{E}+04 \quad 3.75 \mathrm{E}+04 \quad 1.75 \mathrm{E}+04 \quad 1.15 \mathrm{E}+04 \quad 3.30 \mathrm{E}+03 \quad 1.82 \mathrm{E}+03$ $2007 \quad 1.12 \mathrm{E}+05 \quad 6.13 \mathrm{E}+04 \quad 1.87 \mathrm{E}+056.54 \mathrm{E}+043.95 \mathrm{E}+04 \quad 4.82 \mathrm{E}+042.34 \mathrm{E}+049.21 \mathrm{E}+03 \quad 4.79 \mathrm{E}+031.06 \mathrm{E}+03$ $2008 \quad 2.14 \mathrm{E}+05 \quad 8.85 \mathrm{E}+04 \quad 4.65 \mathrm{E}+04 \quad 1.17 \mathrm{E}+05 \quad 3.97 \mathrm{E}+04 \quad 2.51 \mathrm{E}+04 \quad 2.92 \mathrm{E}+04 \quad 1.21 \mathrm{E}+04 \quad 3.80 \mathrm{E}+03 \quad 1.97 \mathrm{E}+03$ $2009 \quad 1.92 \mathrm{E}+05 \quad 1.53 \mathrm{E}+05 \quad 5.53 \mathrm{E}+04 \quad 3.26 \mathrm{E}+04 \quad 7.37 \mathrm{E}+04 \quad 2.37 \mathrm{E}+04 \quad 1.55 \mathrm{E}+04 \quad 1.70 \mathrm{E}+04 \quad 5.66 \mathrm{E}+03 \quad 1.65 \mathrm{E}+03$

Estimated population abundance at 1st Jan 2010
$0.00 \mathrm{E}+00 \quad 1.52 \mathrm{E}+05 \quad 9.07 \mathrm{E}+04 \quad 3.66 \mathrm{E}+04 \quad 2.10 \mathrm{E}+04 \quad 4.57 \mathrm{E}+04 \quad 1.51 \mathrm{E}+04 \quad 1.00 \mathrm{E}+04 \quad 1.03 \mathrm{E}+04 \quad 2.54 \mathrm{E}+03$ Taper weighted geometric mean of the VPA populations:
$1.69 \mathrm{E}+05 \quad 1.06 \mathrm{E}+05 \quad 5.97 \mathrm{E}+043.34 \mathrm{E}+041.88 \mathrm{E}+041.04 \mathrm{E}+04 \quad 5.99 \mathrm{E}+033.50 \mathrm{E}+031.91 \mathrm{E}+031.06 \mathrm{E}+03$ Standard error of the weighted Log(VPA populations) :

| 0.4741 | 0.5573 | 0.6405 | 0.7281 | 0.8218 | 0.9238 | 1.0421 | 1.12 | 1.2188 | 1.3602 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

YEAR $13^{\text {AGE }} 14$

2000 6.81E+01 1.16E+02
2001 4.51E+02 4.86E+01
2002 1.25E+03 3.15E+02
2003 1.95E+03 7.71E+02
2004 1.47E+03 1.22E+03 2005 1.64E+03 8.96E+02 2006 9.55E+02 9.18E+02 2007 5.75E+02 1.18E+02 2008 4.11E+02 2.41E+02 2009 5.60E+02 2.05E+02

Estimated population abundance at 1st Jan 2010
$4.46 \mathrm{E}+022.46 \mathrm{E}+02$
Taper weighted geometric mean of the VPA populations:
$5.94 \mathrm{E}+024.06 \mathrm{E}+02$

Table 5.5.1. Continued
Standard error of the weighted Log(VPA populations) :

$$
1.8109 \quad 2.0915
$$

1
Log catchability residuals.

Fleet : FLT11: Nor trawl rev

| Age |  | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3 | No data for this fleet at this age |  |  |  |  |  |  |  |  |  |
|  | 4 | 0.21 | 1.13 | -0.05 | -0.19 | -0.54 | -0.18 | -0.17 | -0.2 |  |  |
|  | 5 | 0.53 | 0.41 | 0.2 | -0.24 | -0.65 | -0.19 | -0.17 | 0.1 |  |  |
|  | 6 | 0.86 | 0.12 | -0.1 | 0.16 | -0.36 | -0.72 | -0.05 | 0.09 |  |  |
|  | 7 | 0.58 | -0.26 | 0.42 | 0.25 | -0.45 | -0.47 | -0.14 | 0.07 |  |  |
|  | 8 | 0.01 | 0.09 | 0.21 | 0.74 | -0.55 | -0.61 | 0.21 | -0.11 |  |  |
| Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | 3 | No data for this fleet at this age |  |  |  |  |  |  |  |  |  |
|  | 4 | -0.17 | -0.2 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 5 | -0.17 | 0.1 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 6 | -0.05 | 0.09 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 7 | -0.14 | 0.07 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 8 | 0.21 | -0.11 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |

Mean $\log$ catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age | 4 | 5 | 6 | 7 | 8 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Mean Log | -7.8244 | -6.6265 | -5.8016 | -5.4405 | -5.6343 |
| S.E(Log q | 0.4995 | 0.3871 | 0.4549 | 0.3955 | 0.4368 |

Regression statistics:

Ages with $q$ independent of year class strength and constant w.r.t. time.

| Age | Slope |  | t-value |  | Intercept | RSquare | No Pts | Reg s.e |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | Mean Q

Fleet : FLT12: Nor trawl rev

| Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 3 | No data for this fleet at this age |  |  |  |  |  |  |  |  |
| 4 | 99.99 | 99.99 | 0.04 | 0.23 | -1.8 | -0.05 | -0.5 | -0.57 | 1.14 | 1.52 |
| 5 | 99.99 | 99.99 | -0.18 | 0.15 | -0.41 | 0.03 | -0.12 | -0.02 | -0.18 | 0.73 |
| 6 | 99.99 | 99.99 | 0.34 | -0.17 | -0.51 | 0.22 | -0.35 | 0.22 | 0.12 | 0.12 |
| 7 | 99.99 | 99.99 | -0.36 | -0.01 | 0.14 | -0.09 | -0.04 | 0 | 0.34 | 0.02 |
| 8 | 99.99 | 99.99 | -0.12 | 0.4 | 0.3 | -0.15 | -0.1 | 0.01 | 0.06 | -0.39 |

Table 5.5.1. Continued
Mean log catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age | 4 | 5 | 6 | 7 | 8 |
| :---: | ---: | ---: | ---: | ---: | ---: |
| Mean Log | -7.907 | -6.5574 | -5.9819 | -5.6781 | -5.4471 |
| S.E(Log q | 1.0329 | 0.3371 | 0.3039 | 0.1998 | 0.2543 |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.

| Age | Slope |  | t-value |  | Intercept | RSquare | No Pts | Reg s.e |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | Mean Q

Fleet : FLT13: Norway Ac Sur

| Age |  | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3 | -0.48 | -0.36 | 0.34 | -1.05 | 0.18 | 0.28 | 0.38 | 0.7 |  |  |
|  | 4 | -0.41 | -0.22 | -0.02 | 0.78 | 0.67 | -0.06 | 0.02 | -0.76 |  |  |
|  | 5 | -0.36 | -0.07 | -0.09 | 0.11 | 0.83 | 0.63 | -0.92 | -0.12 |  |  |
|  | 6 | -0.01 | -0.42 | -0.47 | 1.09 | 0.91 | 0.3 | -0.38 | -1.03 |  |  |
|  | 7 | 0.01 | -0.11 | -0.65 | 0.1 | 0.04 | 0.47 | 0.24 | -0.09 |  |  |
|  | 8 | No data for | fleet | is age |  |  |  |  |  |  |  |
| Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | 3 | 0.38 | 0.7 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 4 | 0.02 | -0.76 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 5 | -0.92 | -0.12 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 6 | -0.38 | -1.03 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 7 | 0.24 | -0.09 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 | 99.99 |
|  | 8 | No data for | fleet | is age |  |  |  |  |  |  |  |

Mean log catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age | 3 | 4 | 5 | 6 | 7 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Mean Log | -7.1104 | -6.8182 | -7.4546 | -7.6609 | -8.1083 |
| S.E(Log q | 0.576 | 0.5166 | 0.5474 | 0.7282 | 0.3241 |

Table 5.5.1. Continued
Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.

| Age |  | Slope | $t$-value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3 | 1.48 | -0.514 | 4.7 | 0.16 | 8 | 0.9 | -7.11 |
|  | 4 | 1.46 | -0.598 | 4.48 | 0.22 | 8 | 0.79 | -6.82 |
|  | 5 | 1.15 | -0.236 | 6.81 | 0.28 | 8 | 0.68 | -7.45 |
|  | 6 | 0.69 | 0.587 | 8.71 | 0.37 | 8 | 0.53 | -7.66 |
|  | 7 | 0.96 | 0.233 | 8.18 | 0.88 | 8 | 0.34 | -8.11 |
|  | 1 |  |  |  |  |  |  |  |

Fleet : FLT14: Norway Ac Sur

| Age |  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| ---: | :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 3 | 99.99 | 99.99 | -0.09 | -0.1 | 0.62 | -0.18 | -0.25 | 0.34 | -0.36 | 0.03 |
|  | 4 | 99.99 | 99.99 | 0.04 | 0.04 | 0.62 | -0.07 | -0.04 | -0.27 | -0.41 | 0.1 |
|  | 5 | 99.99 | 99.99 | -0.07 | -0.07 | 0.16 | 0 | -0.14 | 0.4 | 0.21 | -0.5 |
|  | 6 | 99.99 | 99.99 | -0.03 | -0.25 | 0.57 | 0.1 | -0.49 | 0.02 | -0.18 | 0.26 |
|  | 7 | 99.99 | 99.99 | -0.44 | 0.03 | 0.01 | 0.2 | 0.04 | 0.31 | -0.02 | -0.13 |

Mean $\log$ catchability and standard error of ages with catchability
independent of year class strength and constant w.r.t. time

| Age | 3 | 4 | 5 | 6 | 7 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Mean Log | -7.0649 | -7.2301 | -8.0519 | -8.6453 | -8.7744 |
| S.E(Log q | 0.3267 | 0.3046 | 0.2698 | 0.3241 | 0.2224 |

Regression statistics :

Ages with $q$ independent of year class strength and constant w.r.t. time.

| Age |  | Slope | t -value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3 | 1.14 | -0.521 | 6.37 | 0.7 | 8 | 0.39 | -7.06 |
|  | 4 | 0.88 | 0.592 | 7.77 | 0.81 | 8 | 0.28 | -7.23 |
|  | 5 | 0.78 | 1.487 | 8.8 | 0.89 | 8 | 0.19 | -8.05 |
|  | 6 | 1.02 | -0.073 | 8.59 | 0.63 | 8 | 0.36 | -8.65 |
|  | 7 | 0.84 | 0.693 | 9.09 | 0.77 | 8 | 0.2 | -8.77 |
|  | 1 |  |  |  |  |  |  |  |

Terminal year survivor and $F$ summaries :
Age 3 Catchability constant w.r.t. time and dependent on age
Year class $=2006$


Table 5.5.1. Continued
$\begin{array}{llll}\text { F shrink: } & 90547 & 1.5 & 0.052\end{array}$
Weighted prediction :

| Survivors | Int | Ext | N |  | Var |
| :---: | :---: | :---: | :---: | :---: | :---: |$\quad$ F

1
Age 4 Catchability constant w.r.t. time and dependent on age
Year class $=2005$

| Fleet |  | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | $N$ | Scaled <br> Weights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FLT11: Nc | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| FLT12: Nc | 413010 | 1.096 | 0 | 0 | 1 | 0.046 | 0.08 |
| FLT13: Nc | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| FLT14: Nc | 81876 | 0.237 | 0.227 | 0.96 | 2 | 0.921 | 0.351 |
| F shrink: | 194472 | 1.5 |  |  |  | 0.034 | 0.163 |
| Weighted prediction : |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |
| at end of ) | s.e | s.e |  | Ratio |  |  |  |
| 90744 | 0.23 | 0.25 | 4 | 1.074 | 0.322 |  |  |

Age 5 Catchability constant w.r.t. time and dependent on age
Year class = 2004

| Fleet |  | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled <br> Weights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FLT11: Nc | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| FLT12: Nc | 77942 | 0.341 | 0.108 | 0.32 | 2 | 0.252 | 0.107 |
| FLT13: Nc | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| FLT14: Nc | 28109 | 0.188 | 0.246 | 1.31 | 3 | 0.732 | 0.272 |
| F shrink: | 41935 | 1.5 |  |  |  | 0.016 | 0.19 |

Weighted prediction :

| Survivors | Int | Ext | N |  | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| at end of ) | s.e | s.e |  |  | Ratio |  |
| 36577 | 0.16 | 0.24 |  | 6 | 1.463 | 0.215 |

Age $6 \begin{array}{ll} & 1 \\ \text { Catchability constant w.r.t. time and dependent on age }\end{array}$
Year class $=2003$

| Fleet | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ |  | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N |  | $\begin{aligned} & \text { aled } \\ & \text { ights } \end{aligned}$ | Estimated <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FLT11: Nc | 1 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| FLT12: Nc | 20357 | 0.235 | 0.128 | 0.54 |  | 3 | 0.342 | 0.251 |
| FLT13: Nc | 1 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| FLT14: Nc | 21253 | 0.164 | 0.142 | 0.87 |  | 4 | 0.646 | 0.241 |
| F shrink: | 22041 | 1.5 |  |  |  |  | 0.011 | 0.234 |

Weighted prediction :

Table 5.5.1. Continued

| Survivors <br> at end of $)$ | s.e | Int | Ext | s.e |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| s.en | Var |  |  |  |  |
| 20951 | 0.13 | 0.09 |  | 8 | 0.636 |

Age 7 Catchability constant w.r.t. time and dependent on age
Year class $=2002$

| Fleet |  | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled <br> Weights | Estimated <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FLT11: Nc | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| FLT12: Nc | 47338 | 0.189 | 0.05 | 0.26 | 4 | 0.42 | 0.27 |
| FLT13: Nc | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| FLT14: Nc | 44477 | 0.152 | 0.111 | 0.73 | 5 | 0.569 | 0.285 |
| F shrink: | 49481 | 1.5 |  |  |  | 0.011 | 0.26 |

Weighted prediction :

| Survivors | Int | Ext | N |  | Var | F |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- |
| at end of J | s.e | s.e |  | Ratio |  |  |
| 45709 | 0.12 | 0.06 |  | 10 | 0.504 | 0.279 |

Age $8 \begin{aligned} & 1 \\ & \text { Catchability constant w.r.t. time and dependent on age }\end{aligned}$
Year class $=2001$


Age 9 Catchability constant w.r.t. time and age (fixed at the value for age) 8
Year class $=2000$

| Fleet | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ |  | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | $N$ |  | led ights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FLT11: Nc | 1 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| FLT12: Nc | 9315 | 0.161 | 0.129 | 0.8 |  | 5 | 0.505 | 0.254 |
| FLT13: Nc | 1 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| FLT14: Nc | 10940 | 0.146 | 0.181 | 1.25 |  | 5 | 0.483 | 0.22 |
| F shrink: | 5334 | 1.5 |  |  |  |  | 0.012 | 0.409 |

Weighted prediction :

| Survivors | Int | Ext | N |  | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| at end of ) | s.e | s.e |  |  | Ratio |  |
| 9998 | 0.11 | 0.1 |  | 11 | 0.954 | 0.239 |

## Table 5.5.1. Continued

Age 10 Catchability constant w.r.t. time and age (fixed at the value for age) 8
Year class $=1999$

| Fleet | Int |  | Ext | Var | N | Scaled |  | Estimated |  |
| :--- | ---: | :--- | :--- | :--- | :--- | :--- | :--- | ---: | :---: |
| FLT11: Nc |  | s.e | s.e | Ratio |  | Weights |  | F |  |
| FLT12: Nc | 10146 | 0.163 | 0.089 | 0.54 | 5 | 0.524 | 0.304 |  |  |
| FLT13: Nc | 1 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
| FLT14: Nc | 10939 | 0.149 | 0.037 | 0.25 | 5 | 0.456 | 0.285 |  |  |
|  |  |  |  |  |  |  |  |  |  |
| F shrink: | 4100 | 1.5 |  |  |  |  | 0.02 | 0.632 |  |

Weighted prediction :

| Survivors | Int | Ext | N |  | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| at end of ) | s.e | s.e |  |  | Ratio |  |
| 10308 | 0.11 | 0.06 |  | 11 | 0.547 | 0.301 |

Age 11 Catchability constant w.r.t. time and age (fixed at the value for age) 8
Year class $=1998$

| Fleet |  | lnt s.e | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var <br> Ratio | N | Scaled <br> Weights | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FLT11: Nc | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| FLT12: Nc | 2182 | 0.16 | 0.102 | 0.64 | 5 | 0.509 | 0.679 |
| FLT13: Nc | 5072 | 0.611 | 0 | 0 | 1 | 0.022 | 0.349 |
| FLT14: Nc | 3014 | 0.16 | 0.133 | 0.83 | 4 | 0.421 | 0.533 |
| F shrink: | 2138 | 1.5 |  |  |  | 0.048 | 0.689 |

Weighted prediction :

| Survivors <br> at end of J | Int | Ext | N |  | Var |
| :---: | :---: | :---: | :---: | :---: | :---: |$\quad \mathrm{F}$.

Age $12 \begin{aligned} & 1 \\ & \text { Catchability constant w.r.t. time and age (fixed at the value for age) } 8\end{aligned}$
Year class $=1997$


Weighted prediction :

| Survivors | Int | Ext | N |  | Var | F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| at end of $)$ | s.e | s.e |  |  | Ratio |  |
| 446 | 0.24 | 0.1 |  | 11 | 0.396 | 1.115 |

Table 5.5.1. Continued
Age 13 Catchability constant w.r.t. time and age (fixed at the value for age) 8
Year class $=1996$

| Fleet | $\begin{aligned} & \text { Int } \\ & \text { s.e } \end{aligned}$ |  | Ext | Var <br> Ratio | N | Scaled |  | Estimated F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | s.e |  |  |  | Weights |  |
| FLT11: Nc | 246 | 0.325 | 0.13 | 0.4 |  | 2 | 0.061 | 0.625 |
| FLT12: Nc | 298 | 0.181 | 0.11 | 0.6 |  | 3 | 0.379 | 0.54 |
| FLT13: Nc | 255 | 0.335 | 0.114 | 0.34 |  | 3 | 0.055 | 0.608 |
| FLT14: Nc | 247 | 0.229 | 0.03 | 0.13 |  | 2 | 0.205 | 0.623 |
| F shrink: | 190 | 1.5 |  |  |  |  | 0.3 | 0.752 |

Weighted prediction :

| Survivors <br> at end of J | Int | Ext | N |  | Var |
| ---: | :---: | :---: | :---: | :---: | :---: |$\quad \mathrm{F}$.

1
Age 14 Catchability constant w.r.t. time and age (fixed at the value for age) 8
Year class = 1995

| Fleet |  | Int | Ext | Var | N |  | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | s.e | s.e | Ratio |  |  | Weights | F |
| FLT11: Nc | 78 | 0.271 | 0.088 | 0.33 |  | 3 | 0.113 | 0.735 |
| FLT12: Nc | 89 | 0.213 | 0.379 | 1.78 |  | 2 | 0.291 | 0.667 |
| FLT13: Nc | 53 | 0.309 | 0.297 | 0.96 |  | 4 | 0.081 | 0.947 |
| FLT14: Nc | 54 | 0.3 | 0 | 0 |  | 1 | 0.13 | 0.938 |
| F shrink: | 106 | 1.5 |  |  |  |  | 0.385 | 0.584 |

Weighted prediction :

| Survivors <br> at end of J | s.e | Int | Ext | N |  | Var |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- |$\quad \mathrm{F}$.



| Table 5.5.2 continue |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Table 8 Fishing mortality ( F ) at age |  |  |  |  |  |  |  |  |  |  |  |  |
|  | YEAF | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0.4509 | 0.376 | 0.1393 | 0.0961 | 0.0382 | 0.0546 | 0.0899 | 0.0816 | 0.0292 | 0.0416 |  |
|  | 4 | 0.5629 | 0.5086 | 0.315 | 0.2921 | 0.1941 | 0.4323 | 0.247 | 0.1368 | 0.1541 | 0.1549 |  |
|  | 5 | 0.4048 | 0.3763 | 0.4395 | 0.5491 | 0.4014 | 0.3878 | 0.2672 | 0.2463 | 0.2019 | 0.3384 |  |
|  | 6 | 0.6026 | 0.3985 | 0.7446 | 0.548 | 0.7932 | 0.4649 | 0.3073 | 0.3207 | 0.371 | 0.2701 |  |
|  | 7 | 0.5998 | 0.4819 | 0.8718 | 0.5779 | 0.6726 | 0.3917 | 0.5505 | 0.3885 | 0.3356 | 0.3951 |  |
|  | 8 | 0.5271 | 0.3939 | 0.9664 | 0.6789 | 0.4997 | 0.3215 | 0.5098 | 0.4383 | 0.2729 | 0.2184 |  |
|  | 9 | 0.1914 | 0.3591 | 0.4494 | 1.0305 | 0.4503 | 0.147 | 0.6258 | 0.2633 | 0.3195 | 0.263 |  |
|  | 10 | 0.3556 | 0.1034 | 0.501 | 0.4745 | 0.9061 | 0.5576 | 0.4021 | 0.1284 | 0.3996 | 0.4564 |  |
|  | 11 | 0.3028 | 0.5433 | 0.2267 | 0.2371 | 0.6759 | 0.2242 | 0.736 | 0.1457 | 0.4119 | 0.4244 |  |
|  | 12 | 3.4434 | 0.2179 | 0.5988 | 0.1244 | 0.5036 | 0.4996 | 1.1191 | 0.4732 | 0.3004 | 0.8619 |  |
|  | 13 | 1.1015 | 13.5155 | 1.2479 | 0.022 | 0.062 | 0.0175 | 2.1325 | 16.8113 | 0.407 | 0.172 |  |
|  | 14 | 0 | 2.9899 | 0 | 0.3802 | 0.5238 | 0.2908 | 1.0146 | 3.6137 | 0 | 0.4388 |  |
|  | +gp | 0 | 2.9899 | 0 | 0.3802 | 0.5238 | 0.2908 | 1.0146 | 3.6137 | 0 | 0.4388 |  |
| 0 | FBAR < | 0.5425 | 0.4413 | 0.5927 | 0.4918 | 0.5153 | 0.4191 | 0.343 | 0.2731 | 0.2657 | 0.2896 |  |
| Table 8 Fishing mortality (F) at age |  |  |  |  |  |  |  |  |  |  |  |  |
|  | YEAF | 2000 | $2001$ | $2002$ | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | FBAR |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0.071 | 0.0213 | 0.021 | 0.0193 | 0.0071 | 0.0749 | 0.0232 | 0.0316 | 0.1394 | 0.0334 | 0.0682 |
|  | 4 | 0.1306 | 0.0811 | 0.1285 | 0.2224 | 0.0644 | 0.1145 | 0.2851 | 0.0771 | 0.2699 | 0.3216 | 0.2229 |
|  | 5 | 0.1388 | 0.2255 | 0.1628 | 0.1395 | 0.2272 | 0.1397 | 0.1575 | 0.269 | 0.1531 | 0.2154 | 0.2125 |
|  | 6 | 0.2609 | 0.2304 | 0.3483 | 0.1426 | 0.154 | 0.2395 | 0.2081 | 0.2991 | 0.2617 | 0.2446 | 0.2685 |
|  | 7 | 0.2538 | 0.2781 | 0.2129 | 0.211 | 0.2193 | 0.2318 | 0.2736 | 0.2521 | 0.3151 | 0.2787 | 0.282 |
|  | 8 | 0.3005 | 0.2 | 0.2681 | 0.3562 | 0.2856 | 0.2565 | 0.2723 | 0.3001 | 0.2841 | 0.2539 | 0.2794 |
|  | 9 | 0.2531 | 0.2767 | 0.2291 | 0.3731 | 0.3087 | 0.4806 | 0.4429 | 0.4566 | 0.3441 | 0.2393 | 0.3467 |
|  | 10 | 0.3446 | 0.3134 | 0.3355 | 0.4044 | 0.6561 | 0.5499 | 0.6774 | 0.6864 | 0.5607 | 0.3013 | 0.5161 |
|  | 11 | 0.3787 | 0.3098 | 0.2538 | 0.4008 | 0.8542 | 0.3032 | 0.936 | 0.6888 | 0.6305 | 0.6113 | 0.6435 |
|  | 12 | 0.2648 | 0.2721 | 0.3864 | 0.522 | 0.76 | 0.3125 | 0.9533 | 0.7479 | 1.058 | 1.1151 | 0.9737 |
|  | 13 | 0.139 | 0.1592 | 0.2849 | 0.272 | 0.2936 | 0.3773 | 1.8876 | 0.6698 | 0.4946 | 0.6275 | 0.5973 |
|  | 14 | 0.2729 | 0.201 | 0.2729 | 0.2791 | 0.668 | 0.5803 | 0.9397 | 0.6646 | 0.684 | 0.6956 | 0.6814 |
|  | +gp | 0.2729 | 0.201 | 0.2729 | 0.2791 | 0.668 | 0.5803 | 0.9397 | 0.6646 | 0.684 | 0.6956 |  |
| 0 | FBAR < | 0.196 | 0.2038 | 0.2131 | 0.1789 | 0.1662 | 0.1814 | 0.2311 | 0.2243 | 0.25 | 0.2651 |  |

Table 5.5.3 Northeast Arctic saithe. Stock number at age


| Table 5.5.3 continue |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Table 10 | Stock number at age (start of year) |  |  |  | Numbers*10**-3 |  | 1996 | 1997 | 1998 | 1999 |  |  |  |
|  | YEAF | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |  |  |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 72448 | 242207 | 379322 | 275236 | 208224 | 357521 | 135132 | 166255 | 118510 | 261608 |  |  |  |
|  | 4 | 43468 | 37788 | 136156 | 270182 | 204703 | 164088 | 277170 | 101127 | 125451 | 94231 |  |  |  |
|  | 5 | 30062 | 20269 | 18605 | 81350 | 165169 | 138035 | 87194 | 177264 | 72211 | 88039 |  |  |  |
|  | 6 | 25000 | 16420 | 11391 | 9815 | 38461 | 90521 | 76689 | 54650 | 113453 | 48312 |  |  |  |
|  | 7 | 17986 | 11205 | 9025 | 4429 | 4646 | 14245 | 46559 | 46177 | 32467 | 64097 |  |  |  |
|  | 8 | 2903 | 8083 | 5665 | 3090 | 2034 | 1941 | 7883 | 21982 | 25635 | 19003 |  |  |  |
|  | 9 | 1015 | 1403 | 4463 | 1765 | 1283 | 1011 | 1152 | 3876 | 11611 | 15976 |  |  |  |
|  | 10 | 414 | 686 | 802 | 2331 | 516 | 670 | 714 | 505 | 2439 | 6907 |  |  |  |
|  | 11 | 635 | 237 | 507 | 398 | 1188 | 171 | 314 | 391 | 363 | 1339 |  |  |  |
|  | 12 | 42 | 384 | 113 | 331 | 257 | 495 | 112 | 123 | 277 | 197 |  |  |  |
|  | 13 | 51 | 1 | 253 | 51 | 239 | 127 | 246 | 30 | 63 | 168 |  |  |  |
|  | 14 | 0 | 14 | 0 | 59 | 41 | 184 | 102 | 24 | 0 | 34 |  |  |  |
|  | +gp | 0 | 9 | 0 | 62 | 24 | 187 | 3 | 9 | 0 | 99 |  |  |  |
| 0 | TOT, | 194022 | 338706 | 566302 | 649100 | 626784 | 769194 | 633270 | 572413 | 502480 | 600010 |  |  |  |
|  | Table 10 | Stock number at age (start of year) |  |  |  | Numbers* $10{ }^{* *}-3$ |  |  |  |  |  |  |  |  |
|  | YEAF | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | GMST | AMST |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 147875 | 200022 | 350218 | 135318 | 157551 | 399644 | 76642 | 111625 | 214493 | 191514 | 0 | 167910 | 187538 |
|  | 4 | 205451 | 112768 | 160311 | 280778 | 108667 | 128085 | 303591 | 61311 | 88546 | 152757 | 151736 | 105316 | 122557 |
|  | 5 | 66078 | 147607 | 85137 | 115424 | 184049 | 83423 | 93518 | 186906 | 46473 | 55346 | 90744 | 60131 | 73039 |
|  | 6 | 51384 | 47090 | 96456 | 59233 | 82196 | 120067 | 59397 | 65411 | 116933 | 32648 | 36577 | 32537 | 41138 |
|  | 7 | 30192 | 32410 | 30620 | 55744 | 42050 | 57693 | 77364 | 39493 | 39709 | 73692 | 20951 | 18034 | 23999 |
|  | 8 | 35350 | 19179 | 20093 | 20263 | 36958 | 27647 | 37464 | 48178 | 25129 | 23723 | 45709 | 10024 | 14285 |
|  | 9 | 12506 | 21429 | 12856 | 12581 | 11619 | 22742 | 17515 | 23362 | 29218 | 15485 | 15091 | 5684 | 8541 |
|  | 10 | 10055 | 7949 | 13304 | 8370 | 7093 | 6986 | 11515 | 9209 | 12115 | 16957 | 9998 | 3304 | 5178 |
|  | 11 | 3583 | 5833 | 4758 | 7788 | 4574 | 3013 | 3300 | 4789 | 3795 | 5662 | 10308 | 1844 | 3078 |
|  | 12 | 717 | 2009 | 3503 | 3022 | 4270 | 1594 | 1822 | 1060 | 1969 | 1654 | 2545 | 1037 | 1899 |
|  | 13 | 68 | 451 | 1253 | 1949 | 1468 | 1635 | 955 | 575 | 411 | 560 | 446 | 396 | 1172 |
|  | 14 | 116 | 49 | 315 | 771 | 1216 | 896 | 918 | 118 | 241 | 205 | 246 | 122 | 776 |
|  | +gp | 60 | 163 | 253 | 501 | 316 | 941 | 494 | 85 | 344 | 224 | 176 |  |  |
| 0 | TOT, | 563435 | 596959 | 779075 | 701742 | 642027 | 854367 | 684494 | 552122 | 579376 | 570427 | 384524 |  |  |

Table 5.5.4 Northeast Arctic saithe. Spawning stock number at age

| Table 11 | Spawning stock number at age (spawning time) |  |  |  |  | Numbers*10**-3 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAF | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 1035 | 634 | 625 | 1252 | 2052 | 681 | 1973 | 916 | 1373 | 1007 |
| 5 | 27404 | 38226 | 22117 | 21299 | 38879 | 56401 | 27811 | 58367 | 29377 | 50934 |
| 6 | 26684 | 21275 | 34749 | 23794 | 21439 | 36914 | 48160 | 23862 | 47324 | 32506 |
| 7 | 25382 | 19412 | 15313 | 23740 | 17443 | 17312 | 24528 | 32625 | 19016 | 35940 |
| 8 | 18298 | 15522 | 14511 | 10087 | 15582 | 10759 | 11493 | 15768 | 22381 | 15040 |
| 9 | 16160 | 12187 | 11468 | 10738 | 7048 | 9841 | 6618 | 8015 | 10379 | 16267 |
| 10 | 8556 | 11701 | 9312 | 8343 | 7608 | 4397 | 5220 | 4471 | 5335 | 7530 |
| 11 | 4457 | 6140 | 9061 | 7121 | 6063 | 4888 | 2874 | 3294 | 2448 | 3747 |
| 12 | 4435 | 3214 | 4540 | 6866 | 5332 | 3804 | 2974 | 1874 | 1961 | 1584 |
| 13 | 1993 | 3303 | 2382 | 3383 | 5199 | 3493 | 2525 | 1867 | 989 | 1348 |
| 14 | 1716 | 1397 | 2603 | 1695 | 2409 | 3749 | 2399 | 1461 | 1051 | 657 |
| +gp | 6218 | 11360 | 7446 | 5902 | 5781 | 4753 | 4457 | 3760 | 2105 | 1227 |
| Table 11 | Spawn | stock n | er at a | spawni | me) | Numbers |  |  |  |  |
| YEAF | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 1698 | 1758 | 1591 | 754 | 1190 | 428 | 760 | 870 | 848 | 598 |
| 5 | 39479 | 44991 | 55531 | 49519 | 23619 | 31612 | 12568 | 17739 | 20894 | 22266 |
| 6 | 51642 | 39123 | 40305 | 51227 | 45583 | 17800 | 25463 | 9274 | 14104 | 15365 |
| 7 | 26546 | 34296 | 30137 | 30375 | 36630 | 24872 | 12259 | 15657 | 6799 | 8674 |
| 8 | 25759 | 17512 | 21694 | 20862 | 20067 | 19579 | 12619 | 6871 | 8889 | 4147 |
| 9 | 11312 | 13847 | 12233 | 15036 | 14144 | 12073 | 8935 | 7180 | 4217 | 5348 |
| 10 | 11796 | 6980 | 8180 | 8582 | 10168 | 8624 | 6840 | 4687 | 4324 | 2473 |
| 11 | 5565 | 6290 | 4122 | 5053 | 5078 | 6052 | 4605 | 3908 | 2905 | 2521 |
| 12 | 2888 | 3558 | 2875 | 2431 | 2969 | 2855 | 2931 | 2700 | 2751 | 1695 |
| 13 | 1172 | 1675 | 1967 | 1742 | 1399 | 1121 | 1036 | 1563 | 1974 | 1541 |
| 14 | 1077 | 665 | 1114 | 1357 | 820 | 754 | 438 | 369 | 1075 | 1319 |
| +gp | 1342 | 1324 | 2256 | 955 | 760 | 686 | 953 | 1000 | 1380 | 1273 |
| Table 11 | Spawn | stock | mber at | (spawni | time) | Numbers* |  |  |  |  |
| YEAF | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 1002 | 563 | 1544 | 671 | 677 | 696 | 747 | 0 | 0 | 0 |
| 5 | 12744 | 27110 | 14219 | 36128 | 17733 | 11921 | 8999 | 6770 | 22068 | 10108 |
| 6 | 14883 | 9126 | 17348 | 7623 | 19145 | 13731 | 12225 | 8974 | 11289 | 39100 |
| 7 | 9730 | 8461 | 5328 | 8859 | 4173 | 7709 | 7361 | 6700 | 4568 | 5493 |
| 8 | 4034 | 5268 | 4706 | 3467 | 4526 | 2407 | 2985 | 3641 | 2206 | 1629 |
| 9 | 2329 | 2107 | 2453 | 2685 | 1714 | 2020 | 1269 | 1224 | 2272 | 701 |
| 10 | 2878 | 1776 | 1383 | 1578 | 1471 | 870 | 859 | 733 | 419 | 984 |
| 11 | 1540 | 1704 | 1195 | 887 | 1056 | 675 | 591 | 307 | 358 | 53 |
| 12 | 1727 | 894 | 1161 | 866 | 532 | 496 | 305 | 373 | 13 | 227 |
| 13 | 1076 | 1007 | 649 | 846 | 592 | 292 | 212 | 191 | 157 | 0 |
| 14 | 1076 | 651 | 692 | 461 | 539 | 373 | 190 | 147 | 17 | 127 |
| +gp | 1200 | 334 | 303 | 870 | 230 | 94 | 10 | 203 | 1 | 70 |

Table 5.5.4 continue

| Table 11 | Spawning stock number at age (spawning time) |  |  |  |  | Numbers*10**-3 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAF | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 0 | 756 | 2723 | 5404 | 4094 | 3282 | 8315 | 3034 | 5018 | 0 |
| 5 | 6012 | 5067 | 5581 | 21151 | 42944 | 30368 | 18311 | 24817 | 5055 | 7043 |
| 6 | 21250 | 13793 | 9454 | 8637 | 32307 | 72417 | 49848 | 24592 | 37439 | 15460 |
| 7 | 17806 | 10980 | 8393 | 4075 | 4181 | 13105 | 42369 | 38327 | 24026 | 47432 |
| 8 | 2903 | 8083 | 5212 | 2750 | 1668 | 1747 | 7331 | 20664 | 23840 | 17483 |
| 9 | 1015 | 1403 | 4017 | 1535 | 1116 | 980 | 1152 | 3605 | 10682 | 14698 |
| 10 | 414 | 686 | 762 | 2075 | 459 | 630 | 714 | 489 | 2341 | 6631 |
| 11 | 635 | 237 | 507 | 398 | 1188 | 171 | 314 | 391 | 363 | 1326 |
| 12 | 42 | 384 | 113 | 324 | 252 | 485 | 112 | 123 | 277 | 191 |
| 13 | 51 | 1 | 253 | 51 | 239 | 127 | 246 | 30 | 63 | 168 |
| 14 | 0 | 14 | 0 | 59 | 41 | 184 | 102 | 24 | 0 | 34 |
| +gp | 0 | 9 | 0 | 62 | 24 | 187 | 3 | 9 | 0 | 99 |
| Table 11 | Spawning stock number at age (spawning time) |  |  |  |  | Numbers*10**-3 |  |  |  |  |
| YEAF | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 0 | 0 | 0 | 0 | 0 | 3843 | 12144 | 3066 | 5313 | 10693 |
| 5 | 5286 | 16237 | 11068 | 16159 | 38650 | 25027 | 37407 | 78501 | 15336 | 16604 |
| 6 | 23637 | 30138 | 75236 | 48571 | 65756 | 98455 | 51081 | 56908 | 98224 | 26445 |
| 7 | 24758 | 30141 | 29089 | 53514 | 40789 | 55962 | 75816 | 38308 | 38121 | 70007 |
| 8 | 33936 | 18604 | 19691 | 19858 | 36588 | 27371 | 37089 | 47214 | 24878 | 23249 |
| 9 | 12255 | 21000 | 12599 | 12330 | 11502 | 22515 | 17515 | 22894 | 28634 | 15175 |
| 10 | 9955 | 7870 | 13171 | 8286 | 7093 | 6986 | 11515 | 8933 | 11752 | 16109 |
| 11 | 3475 | 5658 | 4662 | 7788 | 4574 | 3013 | 3300 | 4645 | 3681 | 5435 |
| 12 | 674 | 1868 | 3363 | 2962 | 4142 | 1562 | 1804 | 1049 | 1949 | 1638 |
| 13 | 68 | 451 | 1253 | 1949 | 1468 | 1635 | 955 | 575 | 411 | 560 |
| 14 | 116 | 49 | 315 | 771 | 1216 | 896 | 918 | 118 | 241 | 205 |
| +gp | 60 | 163 | 253 | 501 | 316 | 941 | 494 | 85 | 344 | 224 |

Table 5.5.5 Northeast Arctic saithe. Stock biomass at age

|  | Table 12 | Stock biomass at age (start of year) |  |  |  | Tonnes |  | 1966 | 1967 | 1968 | 1969 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAF | 1960 | 1961 | 1962 | 1963 | 1964 | 1965 |  |  |  |  |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 65591 | 73969 | 144649 | 218105 | 67629 | 204467 | 99125 | 141366 | 110790 | 206927 |
|  | 4 | 114871 | 70380 | 69332 | 139017 | 227777 | 75591 | 218951 | 101642 | 152366 | 111724 |
|  | 5 | 81216 | 113287 | 65548 | 63122 | 115223 | 167153 | 82420 | 172979 | 87062 | 150951 |
|  | 6 | 73144 | 58319 | 95253 | 65225 | 58767 | 101188 | 132016 | 65411 | 129723 | 89104 |
|  | 7 | 81844 | 62593 | 49376 | 76551 | 56244 | 55822 | 79090 | 105200 | 61317 | 115889 |
|  | 8 | 73740 | 62553 | 58478 | 40649 | 62795 | 43360 | 46318 | 63546 | 90196 | 60610 |
|  | 9 | 78701 | 59350 | 55848 | 52296 | 34326 | 47926 | 32232 | 39031 | 50546 | 79218 |
|  | 10 | 48169 | 65877 | 52425 | 46971 | 42835 | 24757 | 29391 | 25174 | 30035 | 42391 |
|  | 11 | 28701 | 39540 | 58356 | 45858 | 39048 | 31480 | 18512 | 21215 | 15767 | 24131 |
|  | 12 | 31534 | 22848 | 32280 | 48818 | 37913 | 27048 | 21147 | 13323 | 13940 | 11261 |
|  | 13 | 15587 | 25828 | 18629 | 26457 | 40655 | 27312 | 19743 | 14599 | 7731 | 10543 |
|  | 14 | 15304 | 12458 | 23216 | 15121 | 21487 | 33440 | 21398 | 13031 | 9373 | 5864 |
|  | +gp | 59070 | 107924 | 70741 | 56068 | 54923 | 45150 | 42344 | 35722 | 19999 | 11657 |
| 0 | TOTAL | 767473 | 774927 | 794132 | 894257 | 859622 | 884694 | 842688 | 812239 | 778843 | 920271 |
|  | Table 12 | Stock biomass at age (start of year) |  |  |  | Tonnes |  |  |  |  |  |
|  | YEAF | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 186883 | 186451 | 108846 | 152578 | 66085 | 121068 | 181809 | 156621 | 96237 | 146398 |
|  | 4 | 188520 | 195182 | 176643 | 83729 | 132084 | 47543 | 84403 | 96529 | 94075 | 66431 |
|  | 5 | 117001 | 133338 | 164574 | 146756 | 69998 | 93688 | 37246 | 52571 | 61922 | 65988 |
|  | 6 | 141561 | 107244 | 110482 | 140423 | 124952 | 48792 | 69797 | 25421 | 38661 | 42118 |
|  | 7 | 85598 | 110586 | 97176 | 97945 | 118112 | 80198 | 39529 | 50487 | 21922 | 27969 |
|  | 8 | 103808 | 70572 | 87425 | 84076 | 80870 | 78904 | 50854 | 27691 | 35822 | 16713 |
|  | 9 | 55089 | 67436 | 59573 | 73225 | 68880 | 58794 | 43514 | 34966 | 20536 | 26046 |
|  | 10 | 66412 | 39299 | 46054 | 48317 | 57244 | 48552 | 38507 | 26388 | 24343 | 13925 |
|  | 11 | 35837 | 40507 | 26548 | 32543 | 32704 | 38973 | 29655 | 25166 | 18711 | 16237 |
|  | 12 | 20532 | 25297 | 20441 | 17281 | 21110 | 20297 | 20836 | 19195 | 19563 | 12049 |
|  | 13 | 9164 | 13097 | 15386 | 13624 | 10941 | 8764 | 8102 | 12225 | 15438 | 12055 |
|  | 14 | 9604 | 5935 | 9939 | 12101 | 7316 | 6723 | 3907 | 3289 | 9592 | 11770 |
|  | +gp | 12748 | 12577 | 21428 | 9076 | 7225 | 6513 | 9055 | 9500 | 13109 | 12094 |
| 0 | TOTAL | 1032756 | 1007521 | 944517 | 911672 | 797521 | 658808 | 617215 | 540047 | 469932 | 469793 |
|  | Table 12 | Stock biomass at age (start of year) |  |  |  | Tonnes |  |  |  |  |  |
|  | YEAF | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 89484 | 207059 | 93643 | 107989 | 64378 | 74835 | 132805 | 89851 | 49621 | 49598 |
|  | 4 | 127271 | 78779 | 172975 | 89281 | 85329 | 46290 | 45586 | 138598 | 106149 | 55250 |
|  | 5 | 47036 | 101047 | 52221 | 122177 | 65128 | 49353 | 34760 | 32109 | 115634 | 94339 |
|  | 6 | 44649 | 29632 | 53270 | 25110 | 60814 | 39252 | 29912 | 21245 | 28575 | 75650 |
|  | 7 | 32664 | 28490 | 17778 | 36159 | 16523 | 25539 | 21340 | 19899 | 17677 | 16258 |
|  | 8 | 17509 | 23075 | 18400 | 14491 | 20230 | 9532 | 11106 | 14564 | 11871 | 6075 |
|  | 9 | 11997 | 12536 | 11505 | 14309 | 9185 | 9173 | 5455 | 5776 | 13243 | 3238 |
|  | 10 | 16547 | 11348 | 7786 | 8961 | 8917 | 4829 | 4028 | 3986 | 2247 | 4585 |
|  | 11 | 9410 | 11261 | 8582 | 6484 | 6629 | 4647 | 3452 | 1777 | 2480 | 440 |
|  | 12 | 10257 | 6148 | 8374 | 7520 | 3663 | 4037 | 1949 | 2340 | 116 | 1539 |
|  | 13 | 7146 | 6800 | 4545 | 7224 | 4852 | 1772 | 1722 | 1340 | 1235 | 0 |
|  | 14 | 8315 | 4644 | 5555 | 3951 | 4924 | 3605 | 1431 | 1227 | 152 | 1155 |
|  | +gp | 11367 | 2560 | 2860 | 9022 | 1486 | 1288 | 105 | 1725 | 11 | 833 |
| 0 | TOTAL | 433652 | 523380 | 457494 | 452679 | 352057 | 274151 | 293653 | 334438 | 349010 | 308959 |

Table 5.5.5 continue

|  | Table 12 | Stock biomass at age (start of year) |  |  |  | Tonnes |  | 1996 | 1997 | 1998 | 1999 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAF | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 51438 | 164701 | 254146 | 167894 | 108276 | 200212 | 79728 | 103078 | 80587 | 175278 |
|  | 4 | 43468 | 39677 | 137518 | 267480 | 155574 | 129630 | 227279 | 96070 | 125451 | 98942 |
|  | 5 | 43589 | 37499 | 35721 | 134227 | 204810 | 164262 | 115968 | 219808 | 106872 | 127656 |
|  | 6 | 52250 | 39244 | 25971 | 24145 | 81538 | 154791 | 141107 | 93998 | 212156 | 93242 |
|  | 7 | 44785 | 34510 | 25000 | 12622 | 14959 | 40883 | 115466 | 108517 | 83766 | 145500 |
|  | 8 | 10885 | 27078 | 18129 | 9364 | 7792 | 7338 | 29405 | 68146 | 78699 | 56440 |
|  | 9 | 3958 | 6285 | 16649 | 6547 | 6018 | 4103 | 4978 | 16242 | 47955 | 57674 |
|  | 10 | 2788 | 3198 | 5093 | 10468 | 2738 | 3549 | 3814 | 2922 | 13268 | 28319 |
|  | 11 | 3135 | 1334 | 3496 | 2212 | 6722 | 1170 | 1877 | 2648 | 2435 | 6602 |
|  | 12 | 208 | 2418 | 810 | 2169 | 1776 | 3260 | 699 | 815 | 1376 | 1299 |
|  | 13 | 421 | 7 | 1739 | 536 | 1506 | 1002 | 1809 | 218 | 328 | 1262 |
|  | 14 | 0 | 161 | 0 | 400 | 384 | 1685 | 983 | 218 | 0 | 270 |
|  | +gp | 0 | 85 | 0 | 525 | 216 | 1970 | 46 | 99 | 0 | 736 |
| 0 | TOTAL | 256924 | 356196 | 524271 | 638590 | 592309 | 713854 | 723160 | 712779 | 752893 | 793219 |
|  | Table 12 | Stock biomass at age (start of year) |  |  |  | Tonnes |  |  |  |  |  |
|  | YEAF | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 88725 | 150016 | 241650 | 89310 | 110286 | 235790 | 48285 | 81486 | 135130 | 128314 |
|  | 4 | 211614 | 126300 | 161914 | 255508 | 111927 | 113996 | 251981 | 66215 | 86775 | 154284 |
|  | 5 | 107707 | 227315 | 127705 | 163902 | 252147 | 124301 | 133731 | 263538 | 64133 | 93535 |
|  | 6 | 107907 | 96064 | 190019 | 111950 | 156172 | 250940 | 105726 | 121665 | 224511 | 65623 |
|  | 7 | 80613 | 84266 | 77774 | 141589 | 101341 | 124617 | 175615 | 95968 | 91728 | 175387 |
|  | 8 | 110998 | 60222 | 65301 | 52278 | 110135 | 82665 | 102276 | 141642 | 71115 | 63579 |
|  | 9 | 47646 | 77787 | 48466 | 43909 | 39968 | 73685 | 52895 | 78261 | 92329 | 50017 |
|  | 10 | 44344 | 36091 | 57338 | 31388 | 26457 | 26687 | 44908 | 33704 | 41555 | 58163 |
|  | 11 | 20637 | 29457 | 23360 | 32085 | 18935 | 11812 | 13399 | 19968 | 14498 | 19647 |
|  | 12 | 5236 | 11690 | 19935 | 15926 | 21736 | 8192 | 9200 | 5341 | 8052 | 6997 |
|  | 13 | 678 | 2884 | 7755 | 11577 | 8750 | 10235 | 5527 | 3490 | 2066 | 2725 |
|  | 14 | 1222 | 383 | 2379 | 5007 | 7282 | 6058 | 5517 | 619 | 1438 | 1146 |
|  | +gp | 662 | 1768 | 2961 | 5611 | 2498 | 6227 | 4125 | 781 | 2941 | 1639 |
| 0 | TOTAL | 827990 | 904243 | 1026557 | 960039 | 967634 | 1075204 | 953186 | 912680 | 836273 | 821055 |

Table 5.5.6 Northeast Arctic saithe. Spawning stock biomass at age

|  | Table 13 | Spawning stock biomass at age (spawning time) |  |  |  |  | Tonnes 1966 |  | 1967 | 1968 | 1969 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAF | 1960 | 1961 | 1962 | 1963 | 1964 |  |  |  |  |  |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 4 | 1149 | 704 | 693 | 1390 | 2278 | 756 | 2190 | 1016 | 1524 | 1117 |
|  | 5 | 44669 | 62308 | 36051 | 34717 | 63372 | 91934 | 45331 | 95139 | 47884 | 83023 |
|  | 6 | 62173 | 49571 | 80965 | 55441 | 49952 | 86010 | 112214 | 55599 | 110264 | 75739 |
|  | 7 | 80208 | 61341 | 48389 | 75020 | 55119 | 54705 | 77508 | 103096 | 60090 | 113571 |
|  | 8 | 73740 | 62553 | 58478 | 40649 | 62795 | 43360 | 46318 | 63546 | 90196 | 60610 |
|  | 9 | 78701 | 59350 | 55848 | 52296 | 34326 | 47926 | 32232 | 39031 | 50546 | 79218 |
|  | 10 | 48169 | 65877 | 52425 | 46971 | 42835 | 24757 | 29391 | 25174 | 30035 | 42391 |
|  | 11 | 28701 | 39540 | 58356 | 45858 | 39048 | 31480 | 18512 | 21215 | 15767 | 24131 |
|  | 12 | 31534 | 22848 | 32280 | 48818 | 37913 | 27048 | 21147 | 13323 | 13940 | 11261 |
|  | 13 | 15587 | 25828 | 18629 | 26457 | 40655 | 27312 | 19743 | 14599 | 7731 | 10543 |
|  | 14 | 15304 | 12458 | 23216 | 15121 | 21487 | 33440 | 21398 | 13031 | 9373 | 5864 |
|  | +gp | 59070 | 107924 | 70741 | 56068 | 54923 | 45150 | 42344 | 35722 | 19999 | 11657 |
| 0 | TOTSF | 539004 | 570302 | 536072 | 498806 | 504704 | 513878 | 468328 | 480490 | 457349 | 519126 |
|  | Table 13 | Spawning stock biomass at age (spawning time) |  |  |  |  | Tonnes |  |  |  |  |
|  | YEAF | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 4 | 1885 | 1952 | 1766 | 837 | 1321 | 475 | 844 | 965 | 941 | 664 |
|  | 5 | 64351 | 73336 | 90516 | 80716 | 38499 | 51528 | 20485 | 28914 | 34057 | 36294 |
|  | 6 | 120326 | 91157 | 93910 | 119359 | 106209 | 41473 | 59328 | 21608 | 32862 | 35801 |
|  | 7 | 83886 | 108374 | 95233 | 95986 | 115750 | 78594 | 38739 | 49477 | 21484 | 27410 |
|  | 8 | 103808 | 70572 | 87425 | 84076 | 80870 | 78904 | 50854 | 27691 | 35822 | 16713 |
|  | 9 | 55089 | 67436 | 59573 | 73225 | 68880 | 58794 | 43514 | 34966 | 20536 | 26046 |
|  | 10 | 66412 | 39299 | 46054 | 48317 | 57244 | 48552 | 38507 | 26388 | 24343 | 13925 |
|  | 11 | 35837 | 40507 | 26548 | 32543 | 32704 | 38973 | 29655 | 25166 | 18711 | 16237 |
|  | 12 | 20532 | 25297 | 20441 | 17281 | 21110 | 20297 | 20836 | 19195 | 19563 | 12049 |
|  | 13 | 9164 | 13097 | 15386 | 13624 | 10941 | 8764 | 8102 | 12225 | 15438 | 12055 |
|  | 14 | 9604 | 5935 | 9939 | 12101 | 7316 | 6723 | 3907 | 3289 | 9592 | 11770 |
|  | +gp | 12748 | 12577 | 21428 | 9076 | 7225 | 6513 | 9055 | 9500 | 13109 | 12094 |
| 0 | TOTSF | 583641 | 549539 | 568220 | 587140 | 548068 | 439590 | 323825 | 259383 | 246457 | 221057 |
|  | Table 13 | Spawning stock biomass at age (spawning time) |  |  |  |  | Tonnes |  |  |  |  |
|  | YEAF | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 4 | 1273 | 788 | 1730 | 893 | 853 | 926 | 912 | 0 | 0 | 0 |
|  | 5 | 25870 | 55576 | 28722 | 67197 | 35820 | 24677 | 17728 | 11238 | 28909 | 14151 |
|  | 6 | 37952 | 25187 | 45279 | 21343 | 51692 | 36112 | 28118 | 20820 | 27432 | 69598 |
|  | 7 | 32011 | 27920 | 17423 | 35436 | 16192 | 25284 | 21126 | 19899 | 17677 | 16258 |
|  | 8 | 17509 | 23075 | 18400 | 14491 | 20230 | 9532 | 11106 | 14564 | 11871 | 6075 |
|  | 9 | 11997 | 12536 | 11505 | 14309 | 9185 | 9173 | 5455 | 5776 | 13243 | 3238 |
|  | 10 | 16547 | 11348 | 7786 | 8961 | 8917 | 4829 | 4028 | 3986 | 2247 | 4585 |
|  | 11 | 9410 | 11261 | 8582 | 6484 | 6629 | 4647 | 3452 | 1777 | 2480 | 440 |
|  | 12 | 10257 | 6148 | 8374 | 7520 | 3663 | 4037 | 1949 | 2340 | 116 | 1539 |
|  | 13 | 7146 | 6800 | 4545 | 7224 | 4852 | 1772 | 1722 | 1340 | 1235 | 0 |
|  | 14 | 8315 | 4644 | 5555 | 3951 | 4924 | 3605 | 1431 | 1227 | 152 | 1155 |
|  | +gp | 11367 | 2560 | 2860 | 9022 | 1486 | 1288 | 105 | 1725 | 11 | 833 |
|  | TOTSF | 189652 | 187843 | 160760 | 196833 | 164444 | 125880 | 97133 | 84693 | 105371 | 117871 |


| Table 5.5.6 continue |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Table 13 | Spawning stock biomass at age (spawning time) |  |  |  |  | Tonnes1995 | 1996 | 1997 | 1998 | 1999 |
|  | YEAF | 1990 | 1991 | 1992 | 1993 | 1994 |  |  |  |  |  |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 4 | 0 | 794 | 2750 | 5350 | 3111 | 2593 | 6818 | 2882 | 5018 | 0 |
|  | 5 | 8718 | 9375 | 10716 | 34899 | 53250 | 36138 | 24353 | 30773 | 7481 | 10212 |
|  | 6 | 44413 | 32965 | 21556 | 21248 | 68492 | 123833 | 91720 | 42299 | 70012 | 29838 |
|  | 7 | 44337 | 33820 | 23250 | 11613 | 13463 | 37613 | 105074 | 90069 | 61987 | 107670 |
|  | 8 | 10885 | 27078 | 16679 | 8334 | 6389 | 6604 | 27346 | 64057 | 73190 | 51924 |
|  | 9 | 3958 | 6285 | 14984 | 5696 | 5236 | 3980 | 4978 | 15105 | 44119 | 53060 |
|  | 10 | 2788 | 3198 | 4838 | 9317 | 2437 | 3336 | 3814 | 2834 | 12738 | 27186 |
|  | 11 | 3135 | 1334 | 3496 | 2212 | 6722 | 1170 | 1877 | 2648 | 2435 | 6536 |
|  | 12 | 208 | 2418 | 810 | 2126 | 1740 | 3195 | 699 | 815 | 1376 | 1260 |
|  | 13 | 421 | 7 | 1739 | 536 | 1506 | 1002 | 1809 | 218 | 328 | 1262 |
|  | 14 | 0 | 161 | 0 | 400 | 384 | 1685 | 983 | 218 | 0 | 270 |
|  | +gp | 0 | 85 | 0 | 525 | 216 | 1970 | 46 | 99 | 0 | 736 |
| 0 | TOTSF | 118862 | 117519 | 100818 | 102254 | 162947 | 223117 | 269519 | 252018 | 278683 | 289954 |
|  | Table 13 | Spawning stock biomass at age (spawning time) |  |  |  |  | Tonnes |  |  |  |  |
|  | YEAF | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 4 | 0 | 0 | 0 | 0 | 0 | 3420 | 10079 | 3311 | 5207 | 10800 |
|  | 5 | 8617 | 25005 | 16602 | 22946 | 52951 | 37290 | 53492 | 110686 | 21164 | 28060 |
|  | 6 | 49637 | 61481 | 148215 | 91799 | 124937 | 205771 | 90925 | 105849 | 188590 | 53155 |
|  | 7 | 66103 | 78367 | 73885 | 135925 | 98301 | 120878 | 172103 | 93089 | 88059 | 166618 |
|  | 8 | 106558 | 58415 | 63995 | 51232 | 109034 | 81838 | 101253 | 138809 | 70404 | 62307 |
|  | 9 | 46693 | 76231 | 47497 | 43031 | 39568 | 72948 | 52895 | 76696 | 90482 | 49016 |
|  | 10 | 43901 | 35730 | 56765 | 31074 | 26457 | 26687 | 44908 | 32693 | 40309 | 55255 |
|  | 11 | 20018 | 28573 | 22893 | 32085 | 18935 | 11812 | 13399 | 19369 | 14063 | 18861 |
|  | 12 | 4922 | 10872 | 19137 | 15608 | 21084 | 8028 | 9108 | 5288 | 7972 | 6927 |
|  | 13 | 678 | 2884 | 7755 | 11577 | 8750 | 10235 | 5527 | 3490 | 2066 | 2725 |
|  | 14 | 1222 | 383 | 2379 | 5007 | 7282 | 6058 | 5517 | 619 | 1438 | 1146 |
| $\begin{aligned} & \text { +gp } \\ & \text { TOTSF } \end{aligned}$ |  | 662 | 1768 | 2961 | 5611 | 2498 | 6227 | 4125 | 781 | 2941 | 1639 |
|  |  | 349011 | 379708 | 462083 | 445895 | 509797 | 591192 | 563333 | 590680 | 326 | 456509 |

Table 5.5.7 Northeast Arctic saithe. XSA summary
Table 16 Summary (without SOP correction)

Terminal Fs derived using XSA (With F shrinkage)

| Re TOTALE TOTSPE LANDIN YIELD/S FBAR 4-7Age 3 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | 92382 | 767473 | 539004 | 133515 | 0.2477 | 0.3148 |
| 1961 | 104182 | 774927 | 570302 | 105951 | 0.1858 | 0.2421 |
| 1962 | 203732 | 794132 | 536072 | 120707 | 0.2252 | 0.2503 |
| 1963 | 307190 | 894257 | 498806 | 148627 | 0.298 | 0.2737 |
| 1964 | 95252 | 859622 | 504704 | 197426 | 0.3912 | 0.3101 |
| 1965 | 287982 | 884694 | 513878 | 185600 | 0.3612 | 0.268 |
| 1966 | 139613 | 842688 | 468328 | 203788 | 0.4351 | 0.3505 |
| 1967 | 199107 | 812239 | 480490 | 181326 | 0.3774 | 0.2876 |
| 1968 | 156042 | 778843 | 457349 | 110247 | 0.2411 | 0.15 |
| 1969 | 291446 | 920271 | 519126 | 140060 | 0.2698 | 0.1644 |
| 1970 | 263215 | 1032756 | 583641 | 264924 | 0.4539 | 0.3407 |
| 1971 | 262608 | 1007521 | 549539 | 241272 | 0.439 | 0.2954 |
| 1972 | 153304 | 944517 | 568220 | 214334 | 0.3772 | 0.2747 |
| 1973 | 214898 | 911672 | 587140 | 213859 | 0.3642 | 0.2985 |
| 1974 | 93077 | 797521 | 548068 | 264121 | 0.4819 | 0.5102 |
| 1975 | 170518 | 658808 | 439590 | 233453 | 0.5311 | 0.4235 |
| 1976 | 256069 | 617215 | 323825 | 242486 | 0.7488 | 0.5062 |
| 1977 | 220593 | 540047 | 259383 | 182817 | 0.7048 | 0.433 |
| 1978 | 135546 | 469932 | 246457 | 155464 | 0.6308 | 0.4561 |
| 1979 | 206194 | 469793 | 221057 | 164680 | 0.745 | 0.593 |
| 1980 | 113271 | 433652 | 189652 | 144554 | 0.7622 | 0.5049 |
| 1981 | 283643 | 523380 | 187843 | 175540 | 0.9345 | 0.5367 |
| 1982 | 121615 | 457494 | 160760 | 168034 | 1.0452 | 0.5945 |
| 1983 | 102847 | 452679 | 196833 | 156936 | 0.7973 | 0.6101 |
| 1984 | 90673 | 352057 | 164444 | 158786 | 0.9656 | 0.6617 |
| 1985 | 99780 | 274151 | 125880 | 107183 | 0.8515 | 0.5352 |
| 1986 | 225093 | 293653 | 97133 | 67396 | 0.6939 | 0.4729 |
| 1987 | 169531 | 334438 | 84693 | 92391 | 1.0909 | 0.5324 |
| 1988 | 80034 | 349010 | 105371 | 114242 | 1.0842 | 0.5793 |
| 1989 | 67025 | 308959 | 117871 | 122817 | 1.042 | 0.5873 |
| 1990 | 72448 | 256924 | 118862 | 95848 | 0.8064 | 0.5425 |
| 1991 | 242207 | 356196 | 117519 | 107327 | 0.9133 | 0.4413 |
| 1992 | 379322 | 524271 | 100818 | 127604 | 1.2657 | 0.5927 |
| 1993 | 275236 | 638590 | 102254 | 154903 | 1.5149 | 0.4918 |
| 1994 | 208224 | 592309 | 162947 | 146950 | 0.9018 | 0.5153 |
| 1995 | 357521 | 713854 | 223117 | 168378 | 0.7547 | 0.4191 |
| 1996 | 135132 | 723160 | 269519 | 171348 | 0.6358 | 0.343 |
| 1997 | 166255 | 712779 | 252018 | 143629 | 0.5699 | 0.2731 |
| 1998 | 118510 | 752893 | 278683 | 153327 | 0.5502 | 0.2657 |
| 1999 | 261608 | 793219 | 289954 | 150375 | 0.5186 | 0.2896 |
| 2000 | 147875 | 827990 | 349011 | 135928 | 0.3895 | 0.196 |
| 2001 | 200022 | 904243 | 379708 | 135853 | 0.3578 | 0.2038 |
| 2002 | 350218 | 1026557 | 462083 | 154870 | 0.3352 | 0.2131 |
| 2003 | 135318 | 960039 | 445895 | 161592 | 0.3624 | 0.1789 |
| 2004 | 157551 | 967634 | 509797 | 164636 | 0.3229 | 0.1662 |
| 2005 | 399644 | 1075204 | 591192 | 178568 | 0.302 | 0.1814 |
| 2006 | 76642 | 953186 | 563333 | 212822 | 0.3778 | 0.2311 |
| 2007 | 111625 | 912680 | 590680 | 199008 | 0.3369 | 0.2243 |
| 2008 | 214493 | 836273 | 532694 | 184740 | 0.3468 | 0.25 |
| 2009 | 168751 | 798292 | 456509 | 161462 | 0.3537 | 0.2651 |
| ean | 187701 | 697694 | 352841 | 162434 | 0.5939 | 0.3728 |
| its | (Thousar | (Tonnes | (Tonnes | (Tonnes) |  |  |

Table 5.6.1 Northeast Arctic saithe. Yield per recruit
MFYPR version 2 a
Run: y00
Time and date: 10:20 24.04.2010
Yield per results

| FMult | Fbar | CatchNos | Yield | StockNos | Biomass | SpwnNosJan | SSBJan | pwnNosSpw | SSBSpwn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.0000 | 0.0000 | 0.0000 | 0.0000 | 5.5167 | 14.5400 | 3.1238 | 12.0072 | 3.1238 | 12.0072 |
| 0.1000 | 0.0246 | 0.1275 | 0.4268 | 4.8819 | 10.9476 | 2.5199 | 8.4731 | 2.5199 | 8.4731 |
| 0.2000 | 0.0493 | 0.2104 | 0.6123 | 4.4694 | 8.9537 | 2.1365 | 6.5330 | 2.1365 | 6.5330 |
| 0.3000 | 0.0739 | 0.2712 | 0.7054 | 4.1675 | 7.6827 | 1.8623 | 5.3118 | 1.8623 | 5.3118 |
| 0.4000 | 0.0986 | 0.3190 | 0.7576 | 3.9307 | 6.7940 | 1.6518 | 4.4694 | 1.6518 | 4.4694 |
| 0.5000 | 0.1232 | 0.3582 | 0.7893 | 3.7364 | 6.1303 | 1.4826 | 3.8491 | 1.4826 | 3.8491 |
| 0.6000 | 0.1479 | 0.3914 | 0.8100 | 3.5721 | 5.6103 | 1.3423 | 3.3698 | 1.3423 | 3.3698 |
| 0.7000 | 0.1725 | 0.4201 | 0.8239 | 3.4302 | 5.1883 | 1.2234 | 2.9861 | 1.2334 | 2.9861 |
| 0.8000 | 0.1972 | 0.4453 | 0.8335 | 3.3056 | 4.8365 | 1.1210 | 2.6706 | 1.1210 | 2.6706 |
| 0.9000 | 0.2218 | 0.4677 | 0.8401 | 3.1949 | 4.5372 | 1.0316 | 2.4057 | 1.0316 | 2.4057 |
| 1.0000 | 0.2465 | 0.4878 | 0.8445 | 3.0957 | 4.2786 | 0.9528 | 2.1797 | 0.9528 | 2.1797 |
| 1.1000 | 0.2711 | 0.5061 | 0.8473 | 3.0060 | 4.0523 | 0.8829 | 1.9846 | 0.8829 | 1.9846 |
| 1.2000 | 0.2958 | 0.5226 | 0.8489 | 2.9244 | 3.8523 | 0.8204 | 1.8143 | 0.8204 | 1.8143 |
| 1.3000 | 0.3204 | 0.5378 | 0.8494 | 2.8498 | 3.6740 | 0.7642 | 1.6645 | 0.7642 | 1.6645 |
| 1.4000 | 0.3451 | 0.5518 | 0.8492 | 2.7813 | 3.5141 | 0.7135 | 1.5319 | 0.7135 | 1.5319 |
| 1.5000 | 0.3697 | 0.5647 | 0.8484 | 2.7180 | 3.3697 | 0.6676 | 1.4137 | 0.6676 | 1.4137 |
| 1.6000 | 0.3944 | 0.5767 | 0.8472 | 2.6595 | 3.2388 | 0.6258 | 1.3080 | 0.6258 | 1.3080 |
| 1.7000 | 0.4190 | 0.5878 | 0.8455 | 2.6051 | 3.1195 | 0.5877 | 1.2129 | 0.5877 | 1.2129 |
| 1.8000 | 0.4437 | 0.5982 | 0.8435 | 2.5544 | 3.0103 | 0.5528 | 1.1272 | 0.5528 | 1.1272 |
| 1.9000 | 0.4683 | 0.6080 | 0.8413 | 2.5071 | 2.9101 | 0.5208 | 1.0495 | 0.5208 | 1.0495 |
| 2.0000 | 0.4930 | 0.6171 | 0.8390 | 2.4627 | 2.8179 | 0.4913 | 0.9790 | 0.4913 | 0.9790 |


| Reference point |  | : multiplieAbsolute $\mathbf{F}$ |
| :--- | :---: | :---: |
| Fbar(4-7) | 1.0000 | 0.2465 |
| FMax | 1.3210 | 0.3256 |
| F0.1 | 0.3042 | 0.075 |
| F35\%SPR | 0.4396 | 0.1084 |

[^4]Table 5.7.1 Northeast Arctic saithe. Prediction input data
MFDP version 1a
Run: 00
Time and date: 10:01 24.04.2010
Fbar age range: 3-15

| 2010 |  |  |  | M | Mat |  | PF |  | PM |  | SWt |  | Sel | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age |  | N |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 3 |  | 168751 |  | 0.2 |  |  |  |  |  | 0 | 0.677 | 0.0681 | 0.677 |
|  | 4 |  | 133006 |  | 0.2 |  | 0.07 |  |  |  | 0 | 1.023 | 0.2229 | 1.023 |
|  | 5 |  | 90744 |  | 0.2 |  | 0.3 |  |  |  | 0 | 1.493 | 0.2125 | 1.493 |
|  | 6 |  | 36577 |  | 0.2 |  | 0.81 |  |  |  | 0 | 1.930 | 0.2685 | 1.930 |
|  | 7 |  | 20951 |  | 0.2 |  | 0.95 |  |  |  | 0 | 2.373 | 0.282 | 2.373 |
|  | 8 |  | 45709 |  | 0.2 |  | 0.98 |  |  |  | 0 | 2.817 | 0.2794 | 2.817 |
|  | 9 |  | 15091 |  | 0.2 |  | 0.98 |  |  |  | 0 | 3.247 | 0.3467 | 3.247 |
|  | 10 |  | 9998 |  | 0.2 |  | 0.95 |  |  |  | 0 | 3.507 | 0.5161 | 3.507 |
|  | 11 |  | 10308 |  | 0.2 |  | 0.96 |  |  |  | 0 | 3.820 | 0.7382 | 3.820 |
|  | 12 |  | 2545 |  | 0.2 |  | 0.99 |  |  |  | 0 | 4.453 | 0.7382 | 4.453 |
|  | 13 |  | 446 |  | 0.2 |  |  |  |  |  | 0 | 5.323 | 0.7382 | 5.323 |
|  | 14 |  | 246 |  | 0.2 |  |  | 1 |  |  | 0 | 5.597 | 0.7382 | 5.597 |
|  | 15 |  | 176 |  | 0.2 |  |  |  |  |  | 0 | 8.337 | 0.7382 | 8.337 |
| 2011 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Age |  | N |  | M |  | Mat |  | PF |  | PM |  |  | Sel | CWt |
|  | 3 |  | 168751 |  | 0.2 |  |  | 0 |  |  | 0 | 0.677 | 0.0681 | 0.677 |
|  | 4 |  |  |  | 0.2 |  | 0.07 |  |  |  | 0 | 1.023 | 0.2229 | 1.023 |
|  | 5 |  |  |  | 0.2 |  | 0.3 |  |  |  | 0 | 1.493 | 0.2125 | 1.493 |
|  | 6 |  |  |  | 0.2 |  | 0.8 |  |  |  | 0 | 1.930 | 0.2685 | 1.930 |
|  | 7 |  |  |  | 0.2 |  | 0.95 |  |  |  | 0 | 2.373 | 0.282 | 2.373 |
|  | 8 |  |  |  | 0.2 |  | 0.98 |  |  |  | 0 | 2.817 | 0.2794 | 2.817 |
|  | 9 |  |  |  | 0.2 |  | 0.98 |  |  |  | 0 | 3.247 | 0.3467 | 3.247 |
|  | 10 |  |  |  | 0.2 |  | 0.95 |  |  |  | 0 | 3.507 | 0.5161 | 3.507 |
|  | 11 |  |  |  | 0.2 |  | 0.96 |  |  |  | 0 | 3.820 | 0.7382 | 3.820 |
|  | 12 |  |  |  | 0.2 |  | 0.99 |  |  |  | 0 | 4.453 | 0.7382 | 4.453 |
|  | 13 |  |  |  | 0.2 |  |  | 1 |  |  | 0 | 5.323 | 0.7382 | 5.323 |
|  | 14 |  |  |  | 0.2 |  |  | 1 |  |  | 0 | 5.597 | 0.7382 | 5.597 |
|  | 15 |  |  |  | 0.2 |  |  |  |  |  | 0 | 8.337 | 0.7382 | 8.337 |
| 2012 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Age |  | N |  | M |  | Mat |  | PF |  | PM |  |  | Sel | CWt |
|  | 3 |  | 168751 |  | 0.2 |  |  | 0 |  |  | 0 | 0.677 | 0.0681 | 0.677 |
|  | 4 |  |  |  | 0.2 |  | 0.07 |  |  |  | 0 | 1.023 | 0.2229 | 1.023 |
|  | 5 |  |  |  | 0.2 |  | 0.3 |  |  |  | 0 | 1.493 | 0.2125 | 1.493 |
|  | 6 |  |  |  | 0.2 |  | 0.8 |  |  |  | 0 | 1.930 | 0.2685 | 1.930 |
|  | 7 |  |  |  | 0.2 |  | 0.95 |  |  |  | 0 | 2.373 | 0.282 | 2.373 |
|  | 8 |  |  |  | 0.2 |  | 0.98 |  |  |  | 0 | 2.817 | 0.2794 | 2.817 |
|  | 9 |  |  |  | 0.2 |  | 0.98 |  |  |  | 0 | 3.247 | 0.3467 | 3.247 |
|  | 10 |  |  |  | 0.2 |  | 0.95 |  |  |  | 0 | 3.507 | 0.5161 | 3.507 |
|  | 11 |  |  |  | 0.2 |  | 0.96 |  |  |  | 0 | 3.820 | 0.7382 | 3.820 |
|  | 12 |  |  |  | 0.2 |  | 0.99 |  |  |  | 0 | 4.453 | 0.7382 | 4.453 |
|  | 13 |  |  |  | 0.2 |  |  | 1 |  |  | 0 | 5.323 | 0.7382 | 5.323 |
|  | 14 |  |  |  | 0.2 |  |  |  |  |  | 0 | 5.597 | 0.7382 | 5.597 |
|  | 15 |  |  |  | 0.2 |  |  | 1 |  |  | 0 | 8.337 | 0.7382 | 8.337 |

Input units are thousands and kg - output in tonnes

Table 5.7.2 Northeast Arctic saithe. Short term prediction
MFDP version 1a
Run: 00
North-East Arctic saithe
Time and date: 11:30 24.04.2010
Fbar age range: 4-7

| 2010 <br> Biomass | SSB | FMult | FBar | Landings |
| :---: | :---: | :---: | :---: | :---: |
| 774856 | 416334 | 1.3154 | 0.3242 | 204000 |


| 2011 |  |  |  | 2012 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 704195 | 357284 | 0.0000 | 0.0000 | 0 | 860913 | 470829 |
|  | 357284 | 0.1000 | 0.0246 | 16483 | 842519 | 456599 |
|  | 357284 | 0.2000 | 0.0493 | 32429 | 824724 | 442886 |
|  | 357284 | 0.3000 | 0.0739 | 47861 | 807504 | 429668 |
|  | 357284 | 0.4000 | 0.0986 | 62800 | 790834 | 416922 |
|  | 357284 | 0.5000 | 0.1232 | 77267 | 774692 | 404627 |
|  | 357284 | 0.6000 | 0.1479 | 91282 | 759057 | 392764 |
|  | 357284 | 0.7000 | 0.1725 | 104861 | 743908 | 381315 |
|  | 357284 | 0.8000 | 0.1972 | 118023 | 729225 | 370260 |
|  | 357284 | 0.9000 | 0.2218 | 130784 | 714992 | 359585 |
|  | 357284 | 1.0000 | 0.2465 | 143160 | 701189 | 349272 |
|  | 357284 | 1.1000 | 0.2711 | 155166 | 687800 | 339307 |
|  | 357284 | 1.2000 | 0.2958 | 166815 | 674810 | 329675 |
|  | 357284 | 1.3000 | 0.3204 | 178122 | 662204 | 320363 |
|  | 357284 | 1.4000 | 0.3451 | 189100 | 649966 | 311358 |
|  | 357284 | 1.5000 | 0.3697 | 199760 | 638085 | 302648 |
|  | 357284 | 1.6000 | 0.3944 | 210114 | 626545 | 294220 |
|  | 357284 | 1.7000 | 0.4190 | 220174 | 615336 | 286064 |
|  | 357284 | 1.8000 | 0.4437 | 229951 | 604445 | 278170 |
|  | 357284 | 1.9000 | 0.4683 | 239453 | 593861 | 270527 |
|  | 357284 | 2.0000 | 0.4930 | 248692 | 583572 | 263125 |

Input units are thousands and kg - output in tonnes

Table 5.7.3. Northeast Arctic saithe. Short term projection output HCR landings MFDP version 1a
Run: 004
004MFDP Index file 24.04.2010
Time and date: 12:04 24.04.2010
Fbar age range: 4-7

| 2010 |  |  |  |  | 15\% change from 2010 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings |  | 173400 |
| 774856 | 416334 | 1.3154 | 0.3242 | 204000 |  |  |
|  |  |  |  |  |  | average |
| 2011 |  |  |  |  | 2011 | 191275 |
| Biomass | SSB | FMult | FBar | Landings | 2012 | 171076 |
| 704195 | 357284 | 1.2579 | 0.31 | 173400 | 2013 | 156631 |
|  |  |  |  |  |  | 172994 |
| 2012 |  |  |  | 2013 |  |  |
| Biomass | SSB | FMult | FBar | Landings | Biomass SSB |  |
| 667468 | 324248 | 0 | 0 | 0 | 828403 | 440812 |
| . | 324248 | 0.1 | 0.0246 | 15280 | 811206 | 427712 |
| . | 324248 | 0.2 | 0.0493 | 30073 | 794557 | 415080 |
| . | 324248 | 0.3 | 0.0739 | 44398 | 778432 | 402894 |
| . | 324248 | 0.4 | 0.0986 | 58275 | 762812 | 391135 |
| . | 324248 | 0.5 | 0.1232 | 71721 | 747675 | 379785 |
| . | 324248 | 0.6 | 0.1479 | 84754 | 733003 | 368826 |
| . | 324248 | 0.7 | 0.1725 | 97389 | 718777 | 358242 |
| . | 324248 | 0.8 | 0.1972 | 109644 | 704980 | 348018 |
| . | 324248 | 0.9 | 0.2218 | 121531 | 691597 | 338138 |
| . | 324248 | - 1 | 0.2465 | 133066 | 678610 | 328588 |
| . | 324248 | 1.1 | 0.2711 | 144262 | 666005 | 319355 |
| . | 324248 | 1.2 | 0.2958 | 155131 | 653769 | 310426 |
| . | 324248 | 1.3 | 0.3204 | 165687 | 641886 | 301788 |
| . | 324248 | 1.4 | 0.3451 | 175939 | 630345 | 293431 |
| . | 324248 | 1.5 | 0.3697 | 185900 | 619133 | 285344 |
| . | 324248 | 1.6 | 0.3944 | 195580 | 608238 | 277515 |
| . | 324248 | 1.7 | 0.419 | 204990 | 597648 | 269935 |
| . | 324248 | 1.8 | 0.4437 | 214138 | 587354 | 262594 |
| . | 324248 | 1.9 | 0.4683 | 223033 | 577345 | 255484 |
| . | 324248 | 2 | 0.493 | 231686 | 567611 | 248595 |

Input units are thousands and kg - output in tonnes

Table 5.7.4. Northeast Arctic saithe. Detailed short term projection output
MFDP version 1a
Run: 004
Time and date: 12:04 24.04.2010
Fbar age range: 4-7

| Year: <br> Age |  | 2010 F multiplie |  | 1.3154 Fbar: |  | 0.3242 | SSNos(Jar SSB(Jan) |  | SSNos(ST | SSB(ST) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | CatchNos | Yield | StockNos | Biomass |  |  |  |  |
|  | 3 | 0.0896 | 13124 | 8881 | 168751 | 114188 | 0 | 0 | 0 | 0 |
|  | 4 | 0.2932 | 30784 | 31502 | 133006 | 136109 | 9310 | 9528 | 9310 | 9528 |
|  | 5 | 0.2795 | 20149 | 30089 | 90744 | 135511 | 27223 | 40653 | 27223 | 40653 |
|  | 6 | 0.3532 | 9922 | 19150 | 36577 | 70594 | 29627 | 57181 | 29627 | 57181 |
|  | 7 | 0.3709 | 5921 | 14053 | 20951 | 49724 | 19903 | 47238 | 19903 | 47238 |
|  | 8 | 0.3675 | 12819 | 36107 | 45709 | 128747 | 44795 | 126172 | 44795 | 126172 |
|  | 9 | 0.456 | 5047 | 16386 | 15091 | 48995 | 14789 | 48016 | 14789 | 48016 |
|  | 10 | 0.6789 | 4516 | 15836 | 9998 | 35060 | 9498 | 33307 | 9498 | 33307 |
|  | 11 | 0.971 | 5897 | 22528 | 10308 | 39377 | 9896 | 37801 | 9896 | 37801 |
|  | 12 | 0.971 | 1456 | 6484 | 2545 | 11334 | 2520 | 11220 | 2520 | 11220 |
|  | 13 | 0.971 | 255 | 1358 | 446 | 2374 | 446 | 2374 | 446 | 2374 |
|  | 14 | 0.971 | 141 | 788 | 246 | 1377 | 246 | 1377 | 246 | 1377 |
|  | 15 | 0.971 | 101 | 839 | 176 | 1467 | 176 | 1467 | 176 | 1467 |
| Total |  |  | 110132 | 204000 | 534548 | 774856 | 168430 | 416334 | 168430 | 416334 |
| Year: <br> Age |  | 2011 | F multiplie | 1.2579 | Fbar: | 0.31 |  |  |  |  |
|  |  |  | CatchNos | Yield | StockNos | Biomass | SSNos(Jar | SB(Jan) | SSNos(ST | SSB(ST) |
|  | 3 | 0.0857 | 12574 | 8508 | 168751 | 114188 | 0 | 0 | 0 | 0 |
|  | 4 | 0.2804 | 28125 | 28781 | 126323 | 129271 | 8843 | 9049 | 8843 | 9049 |
|  | 5 | 0.2673 | 17344 | 25900 | 81223 | 121293 | 24367 | 36388 | 24367 | 36388 |
|  | 6 | 0.3377 | 14676 | 28324 | 56178 | 108424 | 45504 | 87823 | 45504 | 87823 |
|  | 7 | 0.3547 | 5727 | 13593 | 21036 | 49926 | 19984 | 47430 | 19984 | 47430 |
|  | 8 | 0.3514 | 3198 | 9007 | 11837 | 33342 | 11601 | 32675 | 11601 | 32675 |
|  | 9 | 0.4361 | 8362 | 27148 | 25914 | 84134 | 25396 | 82451 | 25396 | 82451 |
|  | 10 | 0.6492 | 3426 | 12013 | 7831 | 27460 | 7439 | 26087 | 7439 | 26087 |
|  | 11 | 0.9286 | 2311 | 8828 | 4152 | 15860 | 3986 | 15225 | 3986 | 15225 |
|  | 12 | 0.9286 | 1779 | 7922 | 3196 | 14233 | 3164 | 14091 | 3164 | 14091 |
|  | 13 | 0.9286 | 439 | 2338 | 789 | 4201 | 789 | 4201 | 789 | 4201 |
|  | 14 | 0.9286 | 77 | 431 | 138 | 774 | 138 | 774 | 138 | 774 |
|  | 15 | 0.9286 | 73 | 607 | 131 | 1091 | 131 | 1091 | 131 | 1091 |
| Total |  |  | 98110 | 173400 | 507499 | 704195 | 151341 | 357284 | 151341 | 357284 |






Figure 5.1.1 Northeast Arctic saithe (Subareas I and II)




Figure 5.1.1 continued


Figure 5.2.1. Northeast Arctic saithe. Proportion of saithe in the southern half of the survey area (sub area C+D).


Figure 5.4.1 Northeast Arctic saithe. Comparison of SSB and $F_{4-7}$ in 2009 from single fleet and combined XSA runs. SSB and $\mathrm{F}_{4-7}$ in 2008 from an updated 2008-data run is also presented.



Figure 5.5.1. Northeast Arctic saithe. Final run $\log Q$ residuals.


Figure 5.5.2. Northeast arctic saithe. S.E log. Catchability from the four XSA fleet tuning series, final run.



Figure 5.5.3 Northeast Arctic saithe. Estimates of survivors from different fleets and shrinkage and weighting in the final XSA-run.


Figure 5.5.4A. NEA Saithe - Acoustic survey vs. VPA, circle shows last data year.


Figure 5.5.4B. NEA Saithe - Acoustic survey vs. VPA, circle shows last data year.


Figure 5.5.5 Saithe in Sub-areas I and II (Northeast Arctic) RETROSPECTIVE XSA F4-7, recruits and SSB for all fleets.


Fbar

SSB (tons)

Recruitment at age 3 (thousands)

Figure 5.6.1. Northeast Arctic saithe. Fbar, SSB and recruitment for XSA analysis with age span 311+ and

3-15+.


Figure 5.7.1A-B. Northeast arctic saithe. Quantiles of SSB and catch distribution from medium term risk analyses, HCR.

## 6 Beaked redfish (Sebastes mentel/a) in Subareas I and II

ACOM considers any analytical assessments for this stock to be experimental. Until an analytical assessment has been prepared and tested the status of the stock has been deducted from the surveys.

### 6.1 Status of the Fisheries

### 6.1.1 Development of the fishery

A description of the historical development of the fishery in Subareas I and II is found in the Quality handbook for this stock. The Handbook was updated at this year's AFWG.

Since 1 January 2003 the regulations for this stock have been enlarged since from this date all directed trawl fishery for redfish (both S. marinus and S. mentella) outside the permanently closed areas is forbidden in the Norwegian Economic Zone north of $62^{\circ} \mathrm{N}$ and in the Svalbard area. When fishing for other species it is legal to have up to $15 \%$ redfish (both species together) in round weight as bycatch per haul and on board at any time. From 1 January 2006, the maximum bycatch of redfish juveniles in the international shrimp fisheries in the northeast Arctic has been reduced from ten to three redfish per 10 kg shrimp.

A pelagic fishery, for S. mentella, has developed in the Norwegian Sea outside EEZs since 2004 (Figure 6.1-6.2). This fishery, which is further described in Quality handbook for this stock, is managed by the North-East Atlantic Fisheries Commission, and during its $28^{\text {th }}$ annual meeting in November 2009 the Commission adopted by consensus a TAC for 2010 of $8,600 \mathrm{t}$.

### 6.1.2 Bycatch in other fisheries

All catches of $S$. mentella, except the pelagic fishery in the Norwegian Sea outside EEZ, are currently taken as by-catches in other fisheries. Some of the pelagic catches reported on are taken as by-catches in theblue-whiting and herring fisheries.
Numbers and weights of the redfish (fully dominated by S. mentella) taken as bycatch in the shrimp fishery in the Barents Sea during two decades have previously been presented to the AFWG. The results show that shrimp trawlers removed significant numbers of juvenile redfish during the beginning of the 1980's with a peak during 1985 amounting to about 200 millions individuals. As sorting grids became mandatory in 1993, by-catches of redfish reduced drastically during the 1990's. The results also show that closure of areas is necessary to protect the smallest redfish juveniles since these smallest redfish size groups are not sufficiently protected by the sorting grid.

### 6.1.3 Landings prior to 2010 (Tables 6.1-6.5, D1-D2, Figure 6.2)

Nominal catches of S. mentella by country for Sub-areas I and II combined are presented in Table 6.1, and for both redfish species (i.e., S. mentella and S. marinus) in Table D1. The nominal catches by country for Sub-area I and Divisions IIa and IIb are shown in Tables 6.2-6.5. Total international landings in 1965-2009 are also shown in Figure 6.2.

The total landings show a continuous decrease from $48,727 \mathrm{t}$ in 1991 to a historical low at about $8,000 \mathrm{t}$ in 1996 and 1997. Apart from a temporary increase to $18,418 \mathrm{t}$ in 2001, caused by Norwegian trawlers obtaining very good catch rates along the continental slope outside the closed areas in winter 2001, the catches decreased to 2,471 t in 2003 due to stronger regulations enforced.
With the beginning in 2004 of a direct fishery of pelagic redfish in international waters total catches increase considerably. This fishery peaked in 2006 with $28,429 \mathrm{t}$, but has since declined due to the NEAFC regulations. Nevertheless, contrary to the ICES advice of no directed trawl fishery, NEAFC set a TAC of 10,500 t (incl. all by-catches) to be taken in the pelagic trawl fisheries in international waters of the Norwegian Sea in 2009. This was, however, a reduction in TAC from $14,500 \mathrm{t}$ in 2008. According to reports to NEAFC and ICES, only $5,291 \mathrm{t}$ were caught due to generally lower and less profitable catch rates. Not all the countries reported the catches to NEAFC and ICES. EU reported catches are not split by individual country, which is problematic. For this reason catches taken by Spain were recalculated according to the preliminary proportions reported to NEAFC during the fishery.

The redfish population in Sub-area IV (North Sea) is believed to belong to the Northeast Arctic stock. Since this area is outside the traditional areas handled by this Working Group, the catches are not included in the assessment. The total redfish landings from Sub-area IV have up to 2003 been 1,000-3,000 t per year. Since 2004 the annual landings from this area have been about 150-300 t (Table D2).

### 6.1.4 Expected landings in 2010

In 2010 there will be no directed demersal fishery for S. mentella, and all the current regulations will be continued in 2010, including the protection of juveniles from being caught in the shrimp fisheries. Based on the present regulations, the experience from recent years and an increase in the cod and haddock TACs, the total reported demersal by-catches of $S$. mentella for 2010 are expected to be maximum $6,000 \mathrm{t}$.

In addition to this comes, however, the pelagic catches in the Norw egian Sea outside the EEZs. The Northeast Atlantic Fisheries Commission (NEAFC) has set a TAC of $8,600 \mathrm{t}$ for an olympic fishery in these international waters starting 15 August 2010. In total this may lead to landings in 2010 of up to $14,600 \mathrm{t}$.

### 6.2 Data used in the Assessment

No analytical assessment was attempted for this stock this year. All input data sets were, however, updated up to and including 2009.

### 6.2.1 Length-composition from the fishery (Figures 6.3-6.4)

Length distributions of the demersal by-catches of $S$. mentella in the Barents Sea and adjacent waters are shown in Figure 6.3. The main reason for the difference in size compositions between the Portuguese and the rest of the countries was the season of the fishery.
Length compositions from Russia and Portugal of the commercial pelagic catches of S. mentella in the Norwegian Sea outside EEZ in ICES Sub-areas Ila show a different distribution pattern but the same size range (Figure 6.4). This different pattern is difficult to explain, one of the reasons could be the depth or latitude of the fishery.

### 6.2.2 Catch at age (Tables 6.6 and 6.8)

Catch at age for 2008 was revised according to new catch data. Age data for 2009 for demersal S. mentella were available from Norway for all areas, and from Russia in Division Ilb. For the pelagic S.mentella in 2009, age data based on recommended otolith readings were available from Norway (survey) and Russia. Despite the fact that both laboratories base the age reading on otoliths, there are still severe discrepancies in the age readings of $S$. mentella collected in the same area at about the same time. As the difference is related to the ability of reading age of fish of 20 years and more, the problem is believed to be related to the fact that the proximal zone of the otolith sections is not considered by the Russian readers. This problem which also was reported by the ICES Workshop on Age Determination of Redfish (WKADR, ICES 2006) in 2006 must soon be solved through regular otolith exchanges and comparative age readings between international experienced age readers.

Russian total catch-at-length of the demersal fishery in Sub-area I and Division IIa was converted to catch-at-age using the Norwegian age-length keys from Sub-area I and Division IIa (northern part), respectively. The available length distribution from Portuguese catches in Divisions IIa and IIb were converted to catch-at-age using the Norw egian age-length keys from Division IIa (northern part) and Division Ilb. Other countries were assumed to have the same relative age distribution and mean weight as Norway.

Due to uncertainties in the Russian age reading for old fish and potential issues with the length distribution of Portuguese catches in international waters, the catch-at-age figures for 2009 are highly uncertain, in particular for younger (12) and older (18+) age groups. These are presented as preliminary figures.

According to the Norwegian age readings, $77 \%$ of all demersal catches of $S$. mentella are composed of fish older than 18 years. A similar age composition is also seen in the pelagic Norwegian Sea fishery during the survey and beginning of the fishing season in 2009 (Figure 6.14).

### 6.2.3 Weight at age (Tables 6.7 and 6.9)

Catch w eight-at-age data for 2009 w ere available from Norway for all areas, and from Russia from the demersal fishery in Division Ilb and the pelagic fishery. The weight at age in the stock was set equal to the weight at age in the catch. It should be investigated further whether it would be better to use a constant weight-at-age series (e.g., based on survey information) instead of catch weight-at-age which may vary due to changes and selections in the fisheries and not due to growth changes in the stock.

### 6.2.4 Maturity at age (Tables D8a,b)

Age-based maturity ogives for $S$. mentella (sexes combined) were available for the period 1988 to 2001 from Russian research vessel observations in spring (Table D8a). Norw egian data collected in recent years (2004-2008) were used to provide an update of the maturity ogive for the recent period (Table D8b). This indicate an age-at-50\% maturity of 11 y . The detail of the ogive calculation are provided in the report of the NEAFC w orking group on zonal attachment of S. mentella (Anon., 2009b).

### 6.2.5 Scientific surveys (Figures D1 and D2)

The results from the following research vessel survey series were evaluated by the Working Group:
6.2.5.1 Surveys in the Barents Sea and Svalbard area (Tables 1.1, 1.4, D3-D7, Figures 6.5-6.10)

1 ) The international 0-group survey in the Svalbard and Barents Sea areas in August-September, now part of the Ecosystem survey (Table 1.1 and Figures 6.5 and D1).
2 ) Russian bottom trawl survey in the Svalbard and Barents Sea areas in October-December from 1978-2009 in fishing depths of 100-900 m (Table D3, Figures 6.6 and D2F).
3 ) Norwegian Svalbard (Division Ilb) bottom trawl survey (AugustSeptember) from 1986-2009 in fishing depths of 100-500 m (swept area down to 800 m ). Data disaggregated by age only for the years 1992-2009 (Table D4a,b, Figure D2C).
4 ) Norwegian Barents Sea bottom trawl survey (February) from 1986-2010 (joint with Russia since 2000, except 2006 and 2007) in fishing depths of 100-500 m. Data disaggregated by age only for the years 1992-2009 (Tables D5a,b, Figure D2A).

Although the Norwegian Svalbard (August-September) and Barents Sea (February) groundfish surveys are conducted at different times of the year and may overlap in the south of Bear Island area, the two series can be combined to get an approximate total estimate for the whole area by length back to 1986 and by age back to 1992. This has been done in Figures $6.7 \mathrm{a}, \mathrm{b}$.

5 ) The Norwegian survey initially designed for redfish and Greenland halibut is now part of the ecosystem survey and covers the Norwegian Economic Zone (NEZ) and Svalbard incl. north and east of Spitsbergen during August 1996-2009 from less than 100 m to 800 m depth (Table D6, Figures 6.8-6.9 and D2C). This survey includes survey no. 3 above, and has been a joint survey with Russia since 2003, and since then called the Ecosystem survey.
6 ) Russian acoustic survey in April-May from 1992-2001 (except 1994 and 1996) on S. mentella spawning grounds in the western Barents Sea (Table D7).

A considerable reduction in the abundance of 0 -group redfish has been observed since 1991: abundance decreased to only $20 \%$ of the 1979-1990 average. With the exception of an abundance index of twice the 1991-level in 1994, the indices have remained very low. Record low levels of less than $20 \%$ of the 1991-1995 average have been observed for the 1996-1999 year classes. The 2000 year class was stronger than the preceding four year classes. A promising increase was observed since 2005 with the 2007 and 2009 year classes being the strongest observed since 1990, but survey data indicate low abundance of 0-group fish in 2008 (Figure 6.5).
Results from the Ecosystem survey (Table D6 and Figures 6.8-6.9) confirm the stock development as inter preted from the 0-group survey (Figures 6.5), i.e., relative strong 1988-1990 year classes, followed by weaker 1991-1995 year classes, very weak year classes during 1996-2003, and confirming an improved recruitment since then. It also shows how the year classes born before 1991 have grown in biomass. A sudden decrease of S. mentella for ages 9 and older (i.e., larger than about 28 cm ) after 2003 was observed. The WG has earlier reported this decrease as likely related to the increase of $S$. mentella observed in the pelagic fisheries in the Norwegian Sea happening at the same time. This decrease was also seen in Figure 6.4a and b. Some later improvement
in the abundance indices of these year classes may have been caused by fish returning from the pelagic and back to the continental slope. The strong decrease in biomass observed in 2008 from the ecosystem survey was no longer observed in 2009 (Table D6).
Bottom trawl survey estimates for the 2003- and later year classes indicate an improved recruitment (Tables D5, D6, Figure 6.5, 6.7 and 6.9) except 2008 year class. The overall picture of the relative strength of the year classes is similar in the Russian and Norw egian surveys. How ever, both the Russian survey back to 1977 and results from combining the Norwegian Barents Sea February and the Svalbard August surveys back to 1986 (Figure 6.7) show lower and more variable abundance of S. mentella in the 1980s than could be expected from the 0-group indices and when compared with the abundance observed at present.

Figure 6.9 shows that the cod's predation on juvenile ( $5-14 \mathrm{~cm}$ ) redfish during 19862010. This time series confirms the presence of redfish juveniles and may be used as an indicator of small redfish abundance. A clear difference is seen between the abundance/consumption ratio in the 1980s and at present. A change in survey trawl catchability (smaller meshes) from 1993 onwards (Jakobsen et al. 1997) and/or a change in the cod's prey preference may cause this difference. As long as the trawl survey time series has not been corrected for the change in catchability, the abundance index of juvenile redfish during the 1980s might have been considerably higher, if this change in catchability had been corrected for.

The decrease in the abundance of young redfish in the surveys during the 1990s is consistent with the decline in the consumption of redfish by cod (Tables 1.5, 1.6; Figure 6.4 a ). It is important that the estimation of the consumption of redfish by cod is being continued.

Russian acoustic surveys estimating the commercial sized and mature part of the S. mentella stock have been conducted in April-May on the Malangen, Kopytov, and Bear Island Banks since 1992. Until the pelagic surveys in 2007, and with the exception of a trial Norwegian survey between $62-70^{\circ} \mathrm{N}$ in spring 1992, this Russian survey has been the only survey targeting commercial sized S. mentella, though on a limited area of its distribution. The survey has unfortunately not been run since 2001. Table D7 shows a $43 \%$ decrease in the estimated spawning stock biomass from 1992 to 1997 to a low level that was observed up to 2000 inclusive before a three-fold increase in the survey abundance of mature fish was seen in 2001 (Table D7). The strong 1982year class migrating west-southwest and out of the surveyed area could explain this intermediate low level. The next, and to date last year classes contributing significantly to the spawning stock are the 1987-1990 year classes. These are now almost $100 \%$ mature and are likely responsible for the improved recruitment currently seen in the Barents Sea.
6.2.5.2 Surveys along the Norwegian and Barents Seas continental slope (Figures 6.116.12)

A slope survey was carried out by IMR (survey number 2009814) from $18^{\text {th }}$ March to $5^{\text {th }}$ April 2009. The survey was dedicated to the joint study of Sebastes mentella and greater argentine (Argentina silus). The survey included trawling ( 67 stations in total) and hydroacoustics carried out from the commercial trawler "Atlantic Star". For few stations, a multisampler cod-end was used allowing for the collection of trawl samples at 3 different depths, during the same haul. Hydroacoustics was performed at 38 kHz , after standard calibration procedure. Allocation of acoustic energy to different
fish species was done during the scrutinizing, on the basis of trawl catch composition. The equation used for length-dependent target strength of $S$. mentella was $\mathrm{TS}=20 \log (\mathrm{~L})-68$. The survey track and the spatial distribution of $\mathrm{s}_{\mathrm{A}}$ allocated to redfish are illustrated in Figure 6.11. Redfish was found in three regions: 1) between $62^{\circ} \mathrm{N}$ and $63^{\circ} \mathrm{N}$ at bottom depth of $400-700 \mathrm{~m}, 2$ ) between $65^{\circ} 30^{\prime} \mathrm{N}$ and $67^{\circ} \mathrm{N}$ at bottom depth of $400-700 \mathrm{~m}$ and 3 ) between $70^{\circ} \mathrm{N}$ and $74^{\circ} \mathrm{N}$ at bottom depths greater than 400 m . S. mentella tends to distribute in a well defined depth layer, and high concentrations are found between 450 and 650 m , almost independently of the bottom depth (Figure 6.11). High concentrations of beaked redfish can be found along the slope these can locally reach sa values up or above $1000 \mathrm{~m}^{2} / \mathrm{NM}^{2}$, indicating a highly aggregated spatial distribution. This is contrasting with the pelagic summer distribution, which is more evenly spread and where $s_{A}$ values do not generally exceed $100 \mathrm{~m}^{2} / \mathrm{NM}^{2}$.

Age/length distribution: All fish sampled were older than 11y, the maximum recorded age was $53 y$ and mean age was 22.5 y . Males and females have similar age distribution, although females mean length and length-at-age are higher. Fish sampled in the shallow waters ( $<450 \mathrm{~m}$ ) were generally larger and older than the average whilst fish sampled in deeper waters ( $>600 \mathrm{~m}$ and pelagic samples) were generally smaller, but not younger than the average. Size distribution tended to decrease with latitude but this is not true for mean age which was highest at mid-latitude $\left(68^{\circ} \mathrm{N}\right.$ to $70^{\circ} \mathrm{N}$ ). The cumulated length and age distribution are illustrated in Figure 6.12. The mean length $(37.5 \mathrm{~cm})$ and mean age ( 22.5 y ) are consistent with observations from the open Norwegian Sea in summer ( $36.6 \mathrm{~cm}, 25 \mathrm{y}$ ).

### 6.2.5.3 Pelagic surveys in the Norwegian Sea in 2009 (Tables 6.10, Figures 6.13-6.14).

Investigation on the distribution and abundance of redfish in the pelagic Norwegian Sea was coordinated by the ICES Planning Group on Redfish Surveys (ICES, 2009). Unfortunately, among the five expected participants (EU, Faroes, Iceland, Norway and Russia) only Norway was capable of carrying out the survey. The observations were confined to the Northern area of redfish distribution in the Norwegian Sea and the results suffer from serious limitations in area coverage. Despite these limitations, the results from the survey provide confirmation of the observations made in the same area in 2008 and additional work carried out on trawl catchability allow for better abundance estimate. Biological sampling confirms the observations made in 2008 about length (mean length $=36.5 \mathrm{~cm}$ ), age (mean age $=25 y$ ) maturity (all individuals mature) and sex-ratio ( $45 \%$ males, $55 \%$ females) (table 6.10). The vertical distribution is very similar to that observed in the same area in 2008, with maximum concentrations between 400 and 550 m ( $350-550 \mathrm{~m}$ in 2008). This is shallower than what was observed along the slope in spring 2009 ( $450-650 \mathrm{~m}$, see section 6.2 .5 .2 above). The horizontal distribution wasn't extensively analysed but visual inspection of the geographical distribution of $\mathrm{SA}_{\mathrm{A}}$ indicates that only a fraction of the population is located in international waters and this is limited to the Atlantic waters found south of the Mohn Ridge (which crosses at $72-73^{\circ} \mathrm{N}$ ).

As in 2009, an attempt to derive abundance estimates was made, based on both hydroacoustics and trawl catches. The catchability of S. mentella by the Gloria trawl 2048 which was previously assumed to be $100 \%$ (by default) was revised on the basis of recent catchability estimates provided by Bethke et al. (2010). When the same TS equation and catchability coefficients are used for the 2008 and 2009 surveys, the results are highly consistent (Table 6.10). The estimated total biomass is around half a million tonnes. This is likely to be an under-estimate, because the total area covered
by the stock is wider than that covered by the survey. Revision of the target strength equation, which will depend on the result of the ICES w orkshop on hydroacoustics target strength of redfish (WKTAR, June 2010) may alter this estimate, likely upward. The new survey therefore support the results reported to the AFWG in 2009 which indicated a spawning biomass of $500,000 \mathrm{t}$, and this is likely an underestimate (ICES 2009).

### 6.3 Results of the Ass essment

The signals of the various surveys are in agreement. The improved recruitment of 0group and juveniles are confirmed by a couple of surveys from 2007 to 2010, which also confirm lower values of the 2008 year class. It is of vital importance that these younger recruiting year classes be given the strongest possible protection from being taken as by-catch in any fishery, e.g., the shrimp fisheries in the Barents Sea and Svalbard area. This will ensure that they can contribute as much as possible to the stock rebuilding after almost 15 years of very poor recruitment.

It is likely that the strong protection of the last previous good year-classes (i.e., those born before 1991) as these were growing has caused the increased abundance of fish larger than 30 cm seen in both demersal and pelagic surveys (e.g., Figure 6.4).

The WG has previously concluded that any improvement of the stock condition is not expected until a significant increase in spawning stock biomass has been detected in surveys with a following increase in the number of juveniles. Positive signs in that direction are now seen. The only year classes that can contribute to the spawning stock in near future are, however, those prior to 1991 as the following fifteen year classes are very poor. These adult year classes need to be protected as the SSB will continue to be composed mainly from these year classes in the next decade.

### 6.4 Comments to the assessment

Since ACFM/ACOM for many years considered it not necessary to assess this stock every year as long as the status of the stock could be clearly deducted from the demersal surveys, no experimental analytical assessment was attempted. However, in the current context of rapid change in the fisheries dynamics and possible changes in the contribution of the pelagic and demersal components of the stock, management plans and harvesting strategies will suffer from lacking an analytical assessment.

Several European research institutes are currently involved in an EU-project on Management And Monitoring Of Deep-sea Fisheries And Stocks (DEEPFISHMAN) which aims at developing a range of strategy options for the exploitation of deep-sea species in the NE Atlantic. One of the tasks is to develop a GADGET Operating model for S. mentella and to use this to test a suite of possible assessment models. The Gadget model will be based on that developed for S. marinus within this WG, with a single stock split into an immature and mature component, and will be tuned to data surveys and commercial fleets.

The WG finds it appropriate and necessary to conduct a benchmark assessment as a follow up of this project and reiterates its recommendation from last year to hold such a benchmark assessment together with other Sebastes stocks in 2012.

The survey series may still be improved further, and it is imperative for good results that valuable research survey time series are continued, and that Norwegian and Russian research vessels get full access to each other's exclusive economic zones for that purpose. In addition, it is necessary to pursue pelagic surveys in the Norwegian

Sea to cover the whole distribution area, incl. the areas where the bulk of the catches have been taken in recent years. New continental slope surveys may also provide better data to the assessment provided these surveys will continue.

The fact that only $50 \%$ of the allowable NEAFC fish quota of 10500 tonnes in international waters of the Norwegian Sea in 2009 was caught gave some concerns as to what extent this was only caused by an observed effort reduction or was also caused by lower abundance and/or behavioural changes of the fish. A better documentation of the fishing effort involved in the international fishery is therefore highly recommended, and NEAFC is requested to provide such information for future stock assessments and advice.

### 6.5 Biological reference points

Until an analytical assessment is available and used as basis for reference points calculations for this stock, candidate reference points for the biomass could be set at the average biomass level, or at a certain percentage of this level, estimated by the Russian and Norwegian trawl surveys since 1986. ACFM supported these suggestions and stated that U-type reference points could be developed provided that a sufficient long time series demonstrating a dynamic range is available. Also the reference point should be expressed in biomass units (SSB or fishable stock). This should be done before the planned benchmark assessment in 2012. The WG also finds the proposed reference points F0.1 and an appropriate Spawning-stock-per-recruit (SPR) level to be useful reference points for management (see ch. 6.7) and recommends to preparethis for the benchmark assessment. Gadget and other assessment models that eventually will be evaluated during a benchmark assessment should also contribute to the establishment of appropriate reference points.

### 6.6 Management advice

In the Barents Sea and Svalbard area, the stock is still historically low taking all age groups into consideration, and this situation is expected to remain for a considerable period irrespective current management actions. Year-classes recruit to the SSB at old age ( $>10-15$ years old) and surveys indicate failure of recruitment over a long time period. How ever, positive signs in the recruitment have been seen in recent years but it is still uncertain how persistent these might be, as exemplified by the apparent weak year-class in 2008. An estimate of the spawning stock biomass in recent years, based on weight-at-age and maturity-at-age data from Anon (2009b) indicates that this might currently follow and increasing trend $(78,000 t$ in $2000,95,000 t$ in 2001, $99,000 t$ in 2002, $127,000 t$ in 2003, $80,000 \mathrm{t}$ in $2004,75,000 \mathrm{t}$ in $2005,134,000 \mathrm{t}$ in 2006 , $137,000 t$ in 2007, 76,000t in 2008 and 140,000t in 2009). How ever, the large fluctuations in the biomass estimates suggest that the stock is not adequately monitored and that biomass estimates may be highly dependent on fish seasonal migration patterns, accessibility to the survey gear and/or change in the vertical distribution. The protective measures introduced in 2003 should be continued, i.e. the area closures and low bycatch limits should be retained, until a significant increase in the spawning stock biomass (and a subsequent increase in the number of juveniles) has been detected in surveys. Recruitment failure has been observed in surveys for more than a decade. In this connection it is of vital importance that the juvenile age classes be given the strongest protection from being caught as by-catch in any fishery, e.g., the shrimp fisheries in the Barents Sea and Svalbard area. This will ensure that the recruiting year classes can contribute as much as possible to the stock rebuilding.

In the Norwegian Sea, no data is available to describe the historical development of the stock. Results from the pelagic surveys conducted in 2008 and 2009 indicate a possible spawning biomass of at least $500,000 \mathrm{t}$ but such estimate should be handled with caution. Furthermore, it is necessary to preserve this spawning biomass close to the current level since very few new mature individuals will enter the stock for at least the next 12-15 years.

Anticipated increases in TACs for cod and haddock in the Barents Sea will likely result in higher bycatches of redfish. This should be taken into consideration in the management of the stock of S. mentella. High and unreported bycatches in the pelagic trawl fisheries for blue whiting, herring, and mackerel in the Norwegian Sea should be avoided.

The AFWG has earlier estimated the minimum acceptable spawning stock level (MBAL) for S. mentella in ICES Sub-areas I and II to be at least 300000 tons without impairing the recruitment. If this still holds, and how the current SSB is in relation to this is uncertain. It should therefore be the observed recruitment in the Barents Sea that should be decisive when evaluating the spawning and recruitment success. The current size of the mature stock, as estimated from surveys, may at present sustain a small fishery, but will inevitably be reduced in the future due to natural mortality and expected poor new recruitment, and may within some years reach the MBAL level. The poor recruitment in 2008 (after a few years of some promising recruitment) and clear reduction of the biomass in the Barents Sea indicate a need for great caution when monitoring this stock.

The WG considers therefore that the new data (landings and survey) available for this stock do not change the perception of the stock from last year. Therefore, the advice for this fishery in 2011 should be the same as the advice given in 2009 for the 2010 fishery. In order to assess the state of the stock, it is necessary that the whole distribution area of S. mentella in Areas I and II is surveyed, both the pelagic and the demersal components. Coordinated pelagic and demersal surveys should be pursued and particular effort should be put on reducing the uncertainties associated with survey estimates.

A reliable assessment of the stock and proper understanding of the fisheries dynamics are dependent on that complete and detailed catch and landings data from all nations fishing on the resource, as well as accompanying biological data, are provided to ICES and the AFWG.

### 6.7 Implementing the ICES Fmsy framework

During the ICES Workshop on Implementing the ICES Fmsy framework (WKFRAME), the Sebastes mentella stock in Sub-areas I and II was used as a case study (ICES 2010). WKFRAME recommends that the bounds for FMSY proxies should be evaluated in function of the YPR and SPR curves, and that the reproductive capacity of the $S$. mentella stock be at least above $30 \%$ of the SPR at $\mathrm{F}=0$. The YPR curve left of the plateau can be used as low bound (F01 proxy) and a prescribed percent SPR as upper bound. The WKFRAME also illustrates by examples why it is informative and important to carry out sensitivity analyses, particularly assumptions regarding natural mortality, selection pattern, growth (density dependence) and maturity.

The AFWG supports the above recommendation by WKFRAME, and that spawner per recruit curves should be provided. The WG found it premature to adopt the values estimated by WKFRAME directly since the input data, incl. growth parameters
need to be better evaluated before being used for this important purpose. The WG recommends, however, that this should be done as an intersessional work until next years' working group and/or the proposed benchmark assessment in 2012, also including an evaluation of the most appropriate SPR level to be used as reference point for the management of this stock. Evaluations of long lived species with relatively low productivity such as rockfish (Sebastes spp) in the Pacific west coast, concluded that higher SPR values ( $50 \%$ to $60 \%$ ) were required to maintain sustainable exploitation of these stocks (e.g., Dorn 2002).

### 6.8 Response to RGAFNW Technical minutes

It is very unsatisfactory that there are no reference points for this stock in the current rebuilding situation, which, because of its biological characteristics, is very vulnerable. A rebuilding plan should be developed. The AFWG recommend a benchmark assessment to be hold for this stock in 2012 together with other Sebastes stocks.

Table 6.1Sebastes mentella in Sub-areas I and II. Nominal catch (t) by countries in Sub-area I, Divisions IIa and IIb combined.

| Year | Canada | Denmark | Estonia | Faroes | France | Germany | Greenland | Iceland | Ireland |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 |  |  |  | 487 | 85 |  | 23 |  |  |
| 1992 |  |  |  | 23 | 12 |  |  |  |  |
| 1993 | 8 | 4 |  | 13 | 50 | 35 | 1 |  |  |
| 1994 |  | 28 |  | 4 | 74 | 18 | 1 |  | 3 |
| 1995 |  |  |  | 3 | 16 | 176 | 2 |  | 4 |
| 1996 |  |  |  | 4 | 75 | 119 | 3 |  | 2 |
| 1997 |  |  |  | 4 | 37 | 81 | 16 |  | 6 |
| 1998 |  |  |  | 20 | 73 | 100 | 14 |  | 9 |
| 1999 |  |  |  | 73 | 26 | 202 | 50 |  | 3 |
| 2000 |  |  |  | 50 | 12 | 62 | 29 | 48 | 1 |
| 2001 |  |  |  | 74 | 16 | 198 | 17 | 3 | 4 |
| 2002 |  |  | 15 | 75 | 58 | 99 | 18 | 41 | 4 |
| 2003 |  |  |  | 64 | 22 | 32 | 8 | 5 | 5 |
| 2004 |  |  |  | 588 | 13 | 10 | 4 | 10 | 3 |
| 2005 |  |  | 5 | 1147 | 46 | 33 | 39 | 4 | 4 |
| 2006 | 433 |  | 396 | 3808 | 215 | 2483 | 63 | 2513 | 9 |
| 2007 |  |  | 684 | 2197 | 234 | 520 | 29 | 1579 | 6 |
| 2008 |  |  |  | 1849 | 187 | 16 | 25 | 9 | 2 |
| 2009 |  |  |  | 1343 | 15 | 42 | 45 | 63 |  |


| Year | Latvia | Lithuania Netherlands | Norway | Poland | Portugal | Russia | Spain | Sweden | UK E \& W | Uk Scot | EU not split | Sum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 |  |  | 33592 |  | 166 | 14302 | 1 |  | 68 | 3 |  | 50718 |
| 1992 |  |  | 10751 |  | 972 | 3577 | 14 |  | 238 | 3 |  | 17582 |
| 1993 |  |  | 5182 |  | 963 | 6260 | 5 |  | 293 |  |  | 14807 |
| 1994 |  |  | 6511 |  | 895 | 5021 | 30 |  | 124 | 12 |  | 14715 |
| 1995 |  |  | 2646 |  | 927 | 6346 | 67 |  | 93 | 4 |  | 12279 |
| 1996 |  |  | 6053 |  | 467 | 925 | 328 |  | 76 | 23 |  | 10071 |
| 1997 |  |  | 4657 | 1 | 474 | 2972 | 272 |  | 71 | 7 |  | 10595 |
| 1998 |  |  | 9733 | 13 | 125 | 3646 | 177 |  | 93 | 41 |  | 16042 |
| 1999 |  |  | 7884 | 6 | 65 | 2731 | 29 |  | 112 | 28 |  | 13208 |
| 2000 |  |  | 6020 | 2 | 115 | 3519 | 87 |  |  | 130 |  | 12075 |
| 2001 |  |  | 13937 | 5 | 179 | 3775 | 90 |  |  | 120 |  | 20419 |
| 2002 |  |  | 2152 | 8 | 242 | 3904 | 190 |  |  | 188 |  | 8996 |
| 2003 |  |  | 1210 | 7 | 44 | 952 | 47 |  |  | 124 |  | 4523 |
| 2004 |  |  | 1375 | 42 | 235 | 2879 | 257 | 1 |  | 76 |  | 7497 |
| 2005 |  | 7 | 1760 |  | 140 | 5023 | 163 |  |  | 95 |  | 10471 |
| 2006 |  | 845 | 4710 | 2496 | 1804 | 11413 | 710 |  |  | 1027 |  | 34931 |
| 2007 |  | 785 | 3209 | 1081 | 1483 | 5660 | 2181 |  |  | 202 |  | 21857 |
| 2008 | 267 | 117 13 | 2214 | 8 | 713 | 7117 | 463 |  |  | 83 |  | 15091 |
| 2009 |  | 3 | 2766 | 338 | 806 | 3843 | 177 |  |  | 103 | $889^{\circ}$ | 12442 |

* catch not split on countries for EU 2009

Table 6.2Sebastes mentella in Sub-areas I and II. Nominal catch (t) by countries in Sub-area I.

| Year | Faroe <br> Islands | Germany $^{4}$ | Greenland | Norway | Russia $^{5}$ | UK(Eng.\&Wales) | Iceland | Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1991 | - | - | - | 8 | 420 | - | - | 428 |
| 1992 | - |  | - | 561 | 408 | - | - | 969 |
| 1993 | $2^{2}$ | - | - | 16 | 588 | - | - | 606 |
| 1994 | $2^{2}$ | 2 | - | 36 | 308 | - | - | 348 |
| 1995 | $2^{2}$ | - | - | 20 | 203 | - | - | 225 |
| 1996 | - | - | - | 5 | 101 | - | - | 106 |
| 1997 | - | - | $3^{2}$ | 12 | 174 | $1^{2}$ | - | 190 |
| 1998 | $20^{2}$ | - | - | 26 | 378 | - | - | 424 |
| 1999 | $69^{2}$ | - | - | 69 | 489 | - | - | 627 |
| 2000 | - | - | - | 47 | 406 | - | $48^{2}$ | 501 |
| 2001 | - | - | - | 8 | 296 | - | $3^{2}$ | 307 |
| 2002 | - | - | - | 4 | 587 | - | - | 591 |
| 2003 | - | - | - | 6 | 292 | - | - | 298 |
| 2004 | - | - | - | 2 | 355 | - | - | 357 |
| 2005 | - | - | - | $3^{1}$ | 327 | - | - | 330 |
| 2006 | $2^{3}$ | - | - | 12 | 460 | 2 | - | 476 |
| 2007 | - | - | - | 11 | 210 | 20 | - | 241 |
| 2008 | - | - | - | $5^{1}$ | 155 | 2 | - | 162 |
| 20091 | - | - | - | 3 | 80 | - | - | 83 |

${ }^{1}$ Provisional figures.
${ }^{2}$ Split on species according to reports to Norwegian authorities.
${ }^{3}$ Based on preliminary estimates of species breakdown by area.
${ }^{4}$ Includes former GDR prior to 1991.
${ }^{5}$ USSR prior to 1991.

Table 6.3 Sebastes mentella in Sub-areas I and II. Nominal catch ( $\mathbf{t}$ ) by countries in Division IIa (including landings from the pelagic trawl fishery in the international water).

| Year | Estonia | Faroe <br> Islands | France | Germany $^{3}$ | Greenland | Ireland | Norway |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1991 | $487^{2}$ | $72^{2}$ | - | - | - | 32,810 |  |
| 1992 | $23^{2}$ | $7^{2}$ | - | - | - | 9,816 |  |
| 1993 |  | $11^{2}$ | $15^{2}$ | 35 | $1^{2}$ | - | 5,029 |
| 1994 | $2^{2}$ | $33^{2}$ | $16^{2}$ | $1^{2}$ | $2^{2}$ | 6,119 |  |
| 1995 |  | $1^{2}$ | $16^{2}$ | $176^{2}$ | $2^{2}$ | $2^{2}$ | 2,251 |
| 1996 | - | $75^{2}$ | $119^{2}$ | $3^{2}$ | - | 5,895 |  |
| 1997 |  | - | $37^{2}$ | 77 | $12^{2}$ | $2^{2}$ | 4,422 |
| 1998 |  | - | $73^{2}$ | $58^{2}$ | $14^{2}$ | $6^{2}$ | 9,186 |
| 1999 |  | - | $16^{2}$ | $160^{2}$ | $50^{2}$ | $3^{2}$ | 7,358 |
| 2000 |  | $50^{2}$ | $11^{2}$ | $35^{2}$ | $29^{2}$ | - | 5,892 |
| 2001 |  | $63^{2}$ | $12^{2}$ | $161^{2}$ | $17^{2}$ | $4^{2}$ | 13,636 |
| 2002 |  | $37^{2}$ | $54^{2}$ | $59^{2}$ | $18^{2}$ | $4^{2}$ | 1,937 |
| 2003 |  | $58^{2}$ | $18^{2}$ | $17^{2}$ | $8^{2}$ | $5^{2}$ | 1,014 |
| 2004 |  | $555^{2}$ | $8^{2}$ | $4^{2}$ | $4^{2}$ | $3^{2}$ | 987 |
| 2005 |  | $1,101^{2}$ | $36^{2}$ | $17^{2}$ | $38^{2}$ | $4^{2}$ | 1,083 |
| 2006 | 396 | 3,793 | 199 | 2,475 | $52^{2}$ | $8^{2}$ | 3,985 |
| 2007 | 684 | 2,157 | 226 | 519 | $29^{2}$ | $5^{2}$ | 3,043 |
| 2008 | - | $1,821^{6}$ | $179^{2}$ | $9^{2}$ | $24^{2}$ | $2^{2}$ | $1,947^{1}$ |
| 20091 | - | 1,316 | 7 | 23 | 45 | - | 2,300 |

Table 6.3 (Cont'd)

| Year | Sweden | Portugal | Poland | Russia ${ }^{4}$ | Spain | UK <br> (Eng.\& Wales) | UK <br> (Scotland) | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 |  | $159{ }^{2}$ |  | 7,596 | - | $23^{2}$ | - | 41,147 |
| 1992 |  | $824{ }^{2}$ |  | 1,096 | - | $27^{2}$ | - | 11,793 |
| 1993 |  | $648^{2}$ |  | 5,328 | - | $2^{2}$ | - | 11,069 |
| 1994 |  | $687^{2}$ |  | 4,692 | $8^{2}$ | $4^{2}$ | - | 11,564 |
| 1995 |  | $715^{2}$ |  | 5,916 | $65^{2}$ | $41^{2}$ | $2^{2}$ | 9,187 |
| 1996 |  | $429{ }^{2}$ |  | 677 | $5^{2}$ | $42^{2}$ | $19^{2}$ | 7,264 |
| 1997 |  | $410^{2}$ |  | 2,341 | $9{ }^{2}$ | $48^{2}$ | $7^{2}$ | 7,365 |
| 1998 |  | $118^{2}$ |  | 2,626 | $55^{2}$ | $65^{2}$ | $41^{2}$ | 12,242 |
| 1999 |  | $56^{2}$ |  | 1,340 | $14^{2}$ | $94{ }^{2}$ | $26^{2}$ | 9,117 |
| 2000 |  | $98^{2}$ |  | 2,167 | $18^{2}$ | Iceland | 1032,5 | 8,403 |
| 2001 |  | $105^{2}$ |  | 2,716 | $18^{2}$ | - | $95^{25}$ | 16,827 |
| 2002 |  | $124^{2}$ |  | 2,615 | $8^{2}$ | $41^{2}$ | 1572,5 | 5,055 |
| 2003 |  | $17^{2}$ |  | 448 | $8^{2}$ | $5^{2}$ | $102{ }^{2,5}$ | 1,700 |
| 2004 | $1{ }^{2}$ | $86^{2}$ |  | 2,081 | $7{ }^{2}$ | $10^{2}$ | $18^{25}$ | 3,765 |
| 2005 | - | $71^{2}$ |  | 3,307 | $20^{2}$ | $2^{2}$ | 1525 | 5,693 |
| 2006 | Lithu -845 <br> Can-433 | 1,731 | 2,467 | 10,110 | 589 | 2,5132,6 | 9582,5 | 32,895 |
| 2007 | Lithu -785 | 1,395 | 1,079 | 5,061 | 2,159 | 1,579 ${ }^{6}$ | 120,5 | 18,840 |
| 2008 | Lithu -117 | 666 | 1 | 6,442 | 430 | $9^{2}$ | $62^{25}$ | 11,989 |
|  | Latvia - <br> 267 <br> Nether - $13^{2}$ |  |  |  |  |  |  |  |
| 20091 | $E U^{7}-889$ | 764 | 338 | 3,305 | 137 | 63 | 86 | 9,272 |

${ }^{1}$ Provisional figures.
${ }^{2}$ Split on species according to reports to Norwegian authorities.
${ }^{3}$ Includes former GDR prior to 1991.
${ }^{4}$ USSR prior to 1991.
${ }^{5}$ UK(E\&W)+UK(Scot.)
${ }^{6}$ As reported to NEAFC
${ }^{7} \mathrm{EU}$ not split on countries.

Table 6.4 Sebastes mentella in Sub-areas I and II. Nominal catch (t) by countries in Division IIb.

| Year | Canada | Denmark | Faroe <br> Islands | France | Germany ${ }^{4}$ | Greenland | Ireland |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | - | - | - | $13^{2}$ | - | 23 | - |
| 1992 | - | - | - | $5^{2}$ | - | - | - |
| 1993 | $8^{2}$ | $4^{2}$ | - | $35^{2}$ | - | - | - |
| 1994 | - | $28^{2}$ | - | $41^{2}$ | - | - | $1{ }^{2}$ |
| 1995 | - | - | - | - | - | - | $2^{2}$ |
| 1996 | - | - | $4^{2}$ | - | - | - | $2^{2}$ |
| 1997 | - | - | $4^{2}$ | - | 3 | $1{ }^{2}$ | $4^{2}$ |
| 1998 | - | - | - | - | $42^{2}$ | - | $3^{2}$ |
| 1999 | - | - | $4^{2}$ | $10^{2}$ | $42^{2}$ | - | - |
| 2000 | - | - | - | $1{ }^{2}$ | $27^{2}$ | - | $1^{2}$ |
| 2001 | - | - | $11^{2}$ | $4^{2}$ | $37^{2}$ | - | - |
| 2002 | - | - | $38^{2}$ | $4^{2}$ | $40^{2}$ | - | - |
| 2003 | - | - | $6^{2}$ | $4^{2}$ | $15^{2}$ | - | - |
| 2004 | - | - | $33^{2}$ | $5^{2}$ | $6^{2}$ | - | - |
| 2005 | Netherl - $7^{2}$ | Iceland - $2^{2}$ | $46^{2}$ | $10^{2}$ | $17^{2}$ | $1{ }^{2}$ | - |
| 2006 | - | - | $13^{2}$ | $16^{2}$ | $8^{2}$ | $11^{2}$ | $1{ }^{2}$ |
| 2007 | - | - | 40 | $8^{2}$ | 1 | - | $1{ }^{2}$ |
| 2008 | - | - | $28^{2}$ | $8^{2}$ | $7^{2}$ | $1^{2}$ | - |
| $2009{ }^{1}$ | $3^{2}$ | - | $27^{2}$ | $8^{2}$ | $19^{2}$ | - | - |

Table 6.4 (Cont'd)

| Year | Norway | Poland | Portugal | Russia $^{5}$ | Spain | UK(Eng. <br> $\&$ <br> Wales) | UK <br> (Scotland) |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1991 | 774 | - | 7 | 6,286 | 1 | $45^{2}$ | $3^{2}$ | 7,152 |
| 1992 | 374 | - | $148^{2}$ | 2,073 | 14 | $211^{2}$ | $3^{2}$ | 2,828 |
| 1993 | 137 | - | $315^{2}$ | 344 | $57^{3}$ | $291^{2}$ | - | 1,191 |
| 1994 | 356 | - | $208^{2}$ | 21 | $22^{3}$ | $120^{2}$ | $12^{2}$ | 809 |
| 1995 | 375 | - | $212^{2}$ | 227 | $2^{3}$ | $52^{2}$ | $2^{2}$ | 872 |
| 1996 | 153 | - | $38^{2}$ | 147 | $323^{2}$ | $34^{2}$ | $4^{2}$ | 705 |
| 1997 | 223 | $1^{2}$ | $64^{2}$ | 457 | $263^{2}$ | $22^{2}$ | - | 1,042 |
| 1998 | 521 | $13^{2}$ | $7^{2}$ | 642 | $122^{2}$ | $28^{2}$ | $1^{2}$ | 1,379 |
| 1999 | 457 | $6^{2}$ | $9^{2}$ | 902 | $15^{2}$ | $18^{2}$ | $2^{2}$ | 1,465 |
| 2000 | 82 | $2^{2}$ | $17^{2}$ | 946 | $69^{2}$ |  | $27^{26}$ | 1,172 |
| 2001 | 293 | $5^{2}$ | $74^{2}$ | 763 | $72^{2}$ | Estonia | $25^{26}$ | 1,284 |
| 2002 | 210 | $8^{2}$ | $118^{2}$ | 702 | $182^{2}$ | 15 | $31^{26}$ | 1,348 |
| 2003 | 190 | 7 | $27^{2}$ | 212 | $39^{2}$ | - | $22^{26}$ | 522 |
| 2004 | 386 | $42^{2}$ | $149^{2}$ | 443 | $250^{2}$ | - | $58^{26}$ | 1,372 |
| 2005 | 673 | - | $69^{2}$ | 1,389 | $143^{2}$ | 5 | $82^{26}$ | 2,442 |
| 2006 | 688 | 29 | $73^{2}$ | 843 | $121^{2}$ | - | $67^{26}$ | 1,870 |
| 2007 | 155 | 2 | 88 | 389 | $22^{2}$ | - | $62^{26}$ | 769 |
| 2008 | $262^{1}$ | 6 | $47^{2}$ | 520 | $33^{2}$ | - | $19^{26}$ | 931 |
| 20091 | 463 | 1 | $42^{2}$ | 458 | $41^{2}$ | - | $17^{26}$ | 1,079 |

${ }^{1}$ Provisional figures.
${ }^{2}$ Split on species according to reports to Norwegian authorities.
${ }^{3}$ Split on species according to the 1992 catches.
${ }^{4}$ Includes former GDR prior to 1991
${ }^{5}$ USSR prior to 1991.
${ }^{6}$ UK(E\&W)+UK(Scot.)

Table 6.5Sebastes mentella in Sub-areas I and II. Nominal catch (t) by countries of the pelagic fishery in international waters of the Norwegian Sea (see text for further details)

| Year | Can | Estonia | Faroe Islands | France | Germany | Iceland |  | Lithuania |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 |  |  |  |  | 9 |  |  |  |
| 2003 |  |  |  |  | 40 |  |  |  |
| 2004 |  |  | 500 |  | 2 |  |  |  |
| 2005 |  |  | 1,083 |  | 20 |  |  |  |
| 2006 | 433 | 396 | 3,766 | 192 | 2,475 | 2,510 ${ }^{2}$ |  | 845 |
| 2007 | Latvia | 684 | 1,968 ${ }^{2}$ | 226 | 497 | 1,579 ${ }^{2}$ |  | 785 |
| 2008 | 267 | - | 1,797 ${ }^{2}$ | - | - | - |  | 117 |
| 20091 | - | - | 1,253 | - | - | - |  | - |
| Year | Norway | Poland |  | Portugal | Russia | Spain | UK | Total |
| 2002 |  |  |  |  |  |  |  | 9 |
| 2003 |  |  |  |  |  |  |  | 40 |
| 2004 |  |  |  |  | 1,510 |  |  | 2,012 |
| 2005 |  |  |  |  | 3,299 |  |  | 4,402 |
| 2006 | 2,862 | 2,447 |  | 1,697 | 9,390 | 575 | 841 | 28,429 |
| 2007 | 1,813 ${ }^{2}$ | 1,079 |  | 1,377 | 3,645 | 2,155 | - | 15,808 |
| 2008 | $330^{2}$ | - |  | 641 | 4,901 | $390{ }^{1}$ | EU ${ }^{3}$ | 8,443 |
| 20091 | - | 338 |  | 701 | 1,975 | 135 | 889 | 5,291 |

${ }^{1}$ Provisional figures.
${ }^{2}$ As reported to NEAFC
${ }^{3} \mathrm{EU}$ not split on countries.

## Table 6.6. S.mentella in Sub-areas I and II. Catch numbers at age.

| YEAR | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | Catch numbers at age (thous.) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  | 2006 | 2007 | 2008 | $2009{ }^{1}$ |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 6 | 159 | 738 | 662 | 223 | 125 | 37 | 9 | 1 | 117 | 2 | 6 | 11 | 5 | 0 | 0 | 0 | 0 |
| 7 | 159 | 730 | 941 | 634 | 533 | 882 | 83 | 24 | 372 | 40 | 37 | 24 | 44 | 10 | 1 | 0 | 1 |
| 8 | 174 | 722 | 1279 | 1699 | 1287 | 2904 | 441 | 390 | 542 | 252 | 103 | 108 | 128 | 8 | 5 | 1 | 16 |
| 9 | 512 | 992 | 719 | 1554 | 1247 | 4236 | 1511 | 1235 | 976 | 572 | 93 | 148 | 347 | 89 | 32 | 10 | 22 |
| 10 | 2094 | 2561 | 740 | 1236 | 1297 | 3995 | 2250 | 2460 | 925 | 709 | 132 | 427 | 540 | 153 | 52 | 44 | 42 |
| 11 | 3139 | 2734 | 1230 | 1078 | 1244 | 2741 | 3262 | 2149 | 1712 | 532 | 220 | 624 | 567 | 256 | 151 | 128 | 48 |
| 12 | 2631 | 3060 | 2013 | 1146 | 876 | 1877 | 1867 | 1816 | 2651 | 1382 | 384 | 931 | 432 | 877 | 314 | 186 | 1507 |
| 13 | 2308 | 1535 | 4297 | 1413 | 1416 | 1373 | 1454 | 1205 | 2660 | 1893 | 391 | 580 | 1607 | 1980 | 1025 | 492 | 520 |
| 14 | 2987 | 2253 | 3300 | 1865 | 1784 | 1277 | 1447 | 1001 | 1911 | 1617 | 434 | 1385 | 1332 | 2774 | 2466 | 541 | 983 |
| 15 | 1875 | 2182 | 2162 | 880 | 1217 | 1595 | 1557 | 993 | 1773 | 855 | 466 | 1047 | 3174 | 4580 | 2836 | 1444 | 1136 |
| 16 | 1514 | 3336 | 1454 | 621 | 537 | 1117 | 1418 | 932 | 1220 | 629 | 513 | 937 | 1041 | 5154 | 3570 | 1423 | 1623 |
| 17 | 1053 | 1284 | 757 | 498 | 1177 | 784 | 1317 | 505 | 714 | 163 | 199 | 927 | 1216 | 4823 | 4002 | 923 | 1292 |
| 18 | 527 | 734 | 794 | 700 | 342 | 786 | 658 | 596 | 814 | 237 | 231 | 549 | 1024 | 4261 | 2866 | 1730 | 2347 |
| +gp | 6022 | 3257 | 2404 | 2247 | 3568 | 6241 | 3919 | 5705 | 16234 | 4082 | 1193 | 2055 | 4266 | 35350 | 17148 | 16389 | 7389 |
| TOTALNUM | 25154 | 26118 | 22752 | 15794 | 16650 | 29845 | 21193 | 19012 | 32621 | 12965 | 4400 | 9754 | 15725 | 60313 | 34469 | 23311 | 16925 |
| TONSLAND | 12866 | 12721 | 10284 | 8075 | 8597 | 14045 | 11209 | 10075 | 18418 | 6993 | 2520 | 5493 | 8466 | 32895 | 19837 | 13860 | 10434 |

${ }^{1}$ preliminary figures

## Table 6.7. $\quad$ S.mentella in Sub-areas I and II. Catch weights at age (kg).

| YEAR | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 20091 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 6 | 0,13 | 0,19 | 0,17 | 0,16 | 0,14 | 0,2 | 0,18 | 0,14 | 0,15 | 0,1 | 0,11 | 0,13 | 0,09 | 0,13 | 0,13 |  |  |  | 0,21 |
| 7 | 0,18 | 0,22 | 0,23 | 0,22 | 0,16 | 0,2 | 0,21 | 0,19 | 0,22 | 0,15 | 0,15 | 0,17 | 0,14 | 0,17 | 0,17 | 0,14 | 0,14 | 0,29 | 0,20 |
| 8 | 0,21 | 0,26 | 0,25 | 0,24 | 0,19 | 0,25 | 0,25 | 0,23 | 0,22 | 0,22 | 0,20 | 0,22 | 0,22 | 0,22 | 0,21 | 0,23 | 0,25 | 0,30 | 0,35 |
| 9 | 0,27 | 0,28 | 0,28 | 0,3 | 0,21 | 0,31 | 0,29 | 0,29 | 0,28 | 0,26 | 0,25 | 0,29 | 0,28 | 0,27 | 0,28 | 0,29 | 0,33 | 0,30 | 0,43 |
| 10 | 0,34 | 0,31 | 0,33 | 0,34 | 0,28 | 0,42 | 0,33 | 0,33 | 0,33 | 0,31 | 0,30 | 0,34 | 0,33 | 0,33 | 0,34 | 0,34 | 0,19 | 0,32 | 0,43 |
| 11 | 0,35 | 0,33 | 0,38 | 0,37 | 0,32 | 0,44 | 0,38 | 0,38 | 0,37 | 0,36 | 0,34 | 0,38 | 0,39 | 0,38 | 0,38 | 0,42 | 0,33 | 0,36 | 0,47 |
| 12 | 0,42 | 0,38 | 0,44 | 0,4 | 0,37 | 0,47 | 0,46 | 0,43 | 0,44 | 0,42 | 0,39 | 0,43 | 0,43 | 0,43 | 0,43 | 0,45 | 0,30 | 0,49 | 0,52 |
| 13 | 0,46 | 0,46 | 0,47 | 0,44 | 0,41 | 0,59 | 0,48 | 0,48 | 0,49 | 0,44 | 0,44 | 0,44 | 0,45 | 0,43 | 0,45 | 0,46 | 0,29 | 0,43 | 0,54 |
| 14 | 0,51 | 0,43 | 0,5 | 0,45 | 0,47 | 0,67 | 0,51 | 0,54 | 0,53 | 0,51 | 0,48 | 0,52 | 0,50 | 0,50 | 0,50 | 0,49 | 0,48 | 0,63 | 0,55 |
| 15 | 0,58 | 0,43 | 0,57 | 0,49 | 0,53 | 0,69 | 0,55 | 0,59 | 0,56 | 0,56 | 0,53 | 0,56 | 0,54 | 0,54 | 0,55 | 0,53 | 0,48 | 0,56 | 0,62 |
| 16 | 0,59 | 0,45 | 0,58 | 0,55 | 0,58 | 0,71 | 0,6 | 0,61 | 0,62 | 0,62 | 0,59 | 0,57 | 0,59 | 0,58 | 0,56 | 0,54 | 0,51 | 0,55 | 0,62 |
| 17 | 0,58 | 0,52 | 0,62 | 0,58 | 0,66 | 0,74 | 0,66 | 0,64 | 0,66 | 0,63 | 0,62 | 0,60 | 0,57 | 0,61 | 0,59 | 0,55 | 0,61 | 0,64 | 0,64 |
| 18 | 0,59 | 0,57 | 0,65 | 0,67 | 0,71 | 0,74 | 0,65 | 0,66 | 0,67 | 0,67 | 0,65 | 0,59 | 0,62 | 0,64 | 0,61 | 0,56 | 0,59 | 0,32 | 0,65 |
| +gp | 0,7 | 0,67 | 0,66 | 0,79 | 0,81 | 0,85 | 0,79 | 0,75 | 0,81 | 0,77 | 0,70 | 0,73 | 0,75 | 0,72 | 0,70 | 0,66 | 0,68 | 0,64 | 0,67 |

Table 6.8 Pelagic Sebastes mentella in the Norwegian Sea (outside the EEZ). Catch numbers at age.

| Numbers*10**-3 | Age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19+ |
| 2006 | 23 | 93 | 1083 | 323 | 1563 | 3628 | 2514 | 3756 | 29704 |
| 2007 | 75 | 440 | 1331 | 2909 | 3347 | 4138 | 3692 | 3437 | 9114 |
| 2008 | 28 | 146 | 115 | 143 | 214 | 594 | 752 | 753 | 13258 |
| 20091 | 9 | 1314 | 294 | 471 | 889 | 999 | 869 | 1150 | 2981 |

Table $6.9 \quad$ Pelagic Sebastes mentella in the Norwegian Sea (outside the EEZ). Catch weights at age (kg).

|  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| YEAR | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5}$ | $\mathbf{1 6}$ | $\mathbf{1 7}$ | $\mathbf{1 8}$ | $\mathbf{1 9 +}$ |
| 2006 | 0,44 | 0,44 | 0,52 | 0,44 | 0,49 | 0,55 | 0,53 | 0,56 | 0,61 |
| 2007 | 0,39 | 0,43 | 0,41 | 0,48 | 0,50 | 0,52 | 0,55 | 0,57 | 0,64 |
| 2008 | 0,36 | 0,47 | 0,56 | 0,50 | 0,56 | 0,54 | 0,56 | 0,55 | 0,64 |
| $2009^{1}$ | 0,38 | 0,44 | 0,45 | 0,48 | 0,54 | 0,59 | 0,64 | 0,58 | 0,69 |
| ${ }^{1}$ preliminary figures |  |  |  |  |  |  |  |  |  |

Table 6.10: Comparison of results from the Norwegian Sea pelagic surveys in 2008 and 2009.

|  | 2009 | $2008^{1}$ |
| :--- | :--- | :--- |
| mean length (cm) All/M/F ${ }^{2}$ | $36.6 / 36.0 / 37.1$ | $37.0 / 36.4 / 37.5$ |
| mean length (cm) S/DSL/D ${ }^{3}$ | $37.2 / 36.5 / 38.3$ | $37.2 / 36.8 / 39.1$ |
| mean weight (cm) All/M/F | $625 / 609 / 666$ | $619 / 585 / 648$ |
| Mean age (y) All/M/F | $25 / 25 / 24$ | $25 / 25 / 25$ |
| Sex ratio | $45 \%(\mathrm{M}) / 55 \%(\mathrm{~F})$ | $45 \%(\mathrm{M}) / 55 \%(\mathrm{~F})$ |
| Occurrence S. mentella | $100 \%$ | $96 \%$ |
| Catch rates | $3.94 \mathrm{t} / \mathrm{NM}^{2}$ | $3.80 \mathrm{t} / \mathrm{NM}^{2}$ |
| mean sa | $34 \mathrm{~m}^{2} / \mathrm{NM}^{2}$ | $33 \mathrm{~m}^{2} / \mathrm{NM}^{2}$ |
| Total Area | $69,520 \mathrm{NM}^{2}$ | $53,720 \mathrm{NM}^{2}$ |
| Abundance (Acoustics) | $332,000 \mathrm{t}$ | $405,000 \mathrm{t}$ |
| Abundance (Trawl) | $548,000 \mathrm{t}$ |  |

${ }^{1}$ The result for 2008 only concern the northern part of the Norwegian Sea which was surve yed by Norway
${ }^{2} \mathrm{M}=$ males only, $\mathrm{F}=$ females only
${ }^{3}$ S = shallower than DSL, DSL = deep scattering layer, $\mathrm{D}=$ deeper than DSL
${ }^{4}$ The abundance derived from hydroacoustics is calculated assuming a Length-dependent target strength equation of $\mathrm{TS}=20 \log (\mathrm{~L})$ - 68 . The alternative equation $20 \log (\mathrm{~L})-71.3$ would result in abundance estimates raised by a factor of 2
${ }^{5}$ The abundance derived from the trawl catches is corrected for the catchability of redfish by Gloria trawl 2048. This is estimated to be 0.5, from Bethke et al. (2010).


Figure 6.1. Sebastes mentella in Sub-areas I and II. Location of pelagic S. mentella catches by Russian fishing vessels in 2009.


Figure. 6.2. Sebastes mentella in Sub-areas I and II. Total international landings 1965-2009 (thousand tonnes).


Figure 6.3. Sebastes mentella in Sub-areas I and II. Length-distributions of the commercial demersal catches inside EEZ in ICES Sub-areas IIa and IIb by those countries providing length data from their demersal by-catches of S. mentella in 2009.


Figure 6.4. Sebastes mentella in Sub-areas I and II. Length-distributions of the commercial pelagic catches in the Norwegian Sea outside EEZ in ICES Sub-area IIa by those countries providing length data from their pelagic fisheries in 2009.


Figure 6.5. Sebastes mentella in Sub-areas I and II. Abundance indices (in millions) with 95\% confidence limits of 0 -group redfish (believed to be mostly S.mentella) in the international 0-group survey in the Barents Sea and Svalbard areas in August-September 1980-2009, as calculated by the new method, and not corrected for catching efficiency.

Mean catch per hour-trawling of young Sebastes mentella




Figure 6.6 Sebastes mentella in Sub-areas I and II. Catch (numbers of specimens) per hour trawling of different ages of S. mentella in the Russian groundfish survey in the Barents Sea and Svalbard areas (ref. Table D3).



Figure 6.7a. Sebastes mentella in Sub-areas I and II. Abundance indices disaggregated by length when combining the Norwegian bottom trawl surveys 1986-2009 in the Barents Sea (winter) and at Svalbard (summer/fall). Top: absolute index values. Bottom: relative frequencies. Horizontal line indicate the median length in the surveyed population.


Figure 6.7b. Sebastes mentella in Sub-areas I and II. Age disaggregated abundance indices for combined Norwegian bottom trawl surveys 1992-2009 at Svalbard (summer/fall) and in the Barents Sea (winter). Top: absolute numbers. Bottom: rehative frequencies. Vertical black line indicate the start of recording for age 16+ group. Horizontal line indicates the median age in the population ( $50 \%$ frequency).


Figure 6.8. Survey regions and subareas in the ecosystem survey in the Barents Sea and adjacent areas as covered in August-September 2007 by the standard 1800 Campelen research trawl ( 22 mm codend) shallower than about 500 m , and the Alfredo 5 trawl ( 60 mm codend) from 500-1500 m along the continental slope from $68-80^{\circ} \mathrm{N}$. The sub-areas are further depth stratified (ref. Table D6).


Figure 6.9. Abundance of S. mentella during the winter survey (February) in the Barents Sea compared with the consumption of redfish (mainly $S$. mentella) by cod.


Figure 6.10. Sebastes mentella in Sub-areas I and II. Abundance indices (on age) from the Ecosystem survey in August-September 1996-2009 covering the Norwegian Economic Zone (NEZ) and Svalbard incl. the area north and east of Spitsbergen (ref. Table D6). Abundance data in arctic waters are not included for 2009.


Figure 6.11. Sebastes mentella in Sub-areas I and II. Horizontal and vertical distribution of S.mentella hydroacoustic backscaterring ( $\mathrm{s}_{\mathrm{A}}$ ) during the Norwegian slope surve y in spring 2009. On the top-left panel, circles are proportional to the $s_{A}$ assigned to redfish along the vessel track. The top-right panel shows the distribution of mean $s_{A}$ by depth and latitude strata (dark blue $=$ no data). The bottom panel shows the vertical distribution of median $s_{A}$ as a function of bottom depth, revealing a preferred depth range for $S$. mentella of $450-650 \mathrm{~m}$ and dominance of pelagic $v \mathrm{vs}$. demersal distributions.


Figure 6.12. Sebastes mentella in Sub-areas I and II. Cumulated distribution of length (left) and age (right) of $S$. mentella as a function of sex (top), depth (middle) and latitude (bottom.)


Figure 6.13. Sebastes mentella in Sub-areas I and II. Left: Spatial distribution of area backscattering coefficient ( $\mathrm{s}_{\mathrm{A}}$ ) of $S$. mentella $\left(\mathrm{m}^{2} / \mathrm{NM}^{2}\right)$ during the Norwegian Sea pelagic survey in summer 2009. Right: cumulated density distribution of catch rates (black) and area backscattering coefficient ( $\mathrm{s}_{\mathrm{A}}$, red) as a function of depth. Dotted lines indicate the 5 and $95 \%$ probability levels. Dashed lines indicate the $\mathbf{2 5 \%}$ and $75 \%$ probabilit y levels.


Figure 6.14. Sebastes mentella in Sub-areas I and II. Age distribution of S. mentella sampled during the open Norwegian Sea survey in July-August 2009. Dots show the proportion at age for individual age. The black line is a smooth fit, which is belie ved to be more reliable when precision in age reading is uncertain. The sampled population is dominated by individuals of 16 years and over with $20-22$ y dominating. Only a fraction ( $40 \%$ ) of otoliths collected during the survey were read at the time of reporting. The estimated smoothed age distribution in the same area in 2008 is indicated as a red dotted line.

Table D1 REDFISH in Sub-areas I and II. Nominal catch (t) by countries in Sub-area I, Divisions IIa and IIb combined as officially reported to ICES.

| Year Can ada | Den Faroe mark Islands |  | Franc | Ger many | Gree <br> ${ }^{4}$ land | nIce land | Ire <br> land | Nethe <br> dlands | rNor P <br> way | Po <br> land | Port ugal | Russia5 | ${ }^{5}$ Spain | nUK (E\& | UK <br> (Scot.) | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984- | - | - | 2,970 | 7,457 | - | - | - | - | 18,650 | - | 1,806 | 669,689 | 25 | 716 | - | 101,313 |
| 1985 - | - | - | 3,326 | 6,566 | - | - | - | - | 20,456 |  | 2,056 | 659,943 | 38 | 167 | - | 92,552 |
| 1986 - | - | 29 | 2,719 | 4,884 | - | - | - | - | 23,255 |  | 1,591 | 20,694 | - | 129 | 14 | 53,315 |
| 1987 - | + | $450{ }^{3}$ | 1,611 | 5,829 | - | - | - | - | 18,051 | - | 1,175 | 57,215 | 25 | 230 | 9 | 34,595 |
| 1988 - | - | 973 | 3,349 | 2,355 | - | - | - | - | 24,662 |  | 500 | 9,139 | 26 | 468 | 2 | 41,494 |
| 1989 - | - | 338 | 1,849 | 4,245 | - | - | - | - | 25,295 | - | 340 | 14,344 | $5{ }^{2}$ | 271 | 1 | 46,688 |
| 1990 - | 373 | 386 | 1,821 | 6,741 | - | - | - | - | 34,090 | - | 830 | 18,918 | - | 333 | - | 63,156 |
| 1991 - | 23 | 639 | 791 | 981 | - | - | - | - | 49,463 |  | 166 | 15,354 | 1 | 336 | 13 | 67,768 |
| 1992 - | 9 | 58 | 1,301 | 530 | 614 | - | - | - | 23,451 | - | 977 | 4,335 | 16 | 479 | 3 | 31,773 |
| $19938^{3}$ | 4 | 152 | 921 | 685 | 15 | - | - | - | 18,319 | - | 1,0 | 7,573 | 65 | 734 | 1 | 29,517 |
| 1994 - | 28 | 26 | 771 | 1026 | 6 | 4 | 3 | - | 21,466 |  | 985 | 6,220 | 34 | 259 | 13 | 30,841 |
| 1995 - | - | 30 | 748 | 692 | 7 | 1 | 5 | 1 | 16,162 |  | 936 | 6,985 | 67 | 252 | 13 | 25,899 |
| 1996 - | - | $42^{3}$ | 746 | 618 | 37 | - | 2 | - | 21,675 |  | 523 | 1,641 | 408 | 305 | 121 | 26,118 |
| 1997 - | - | 7 | 1,011 | 538 | $39^{2}$ | - | 11 | - | 18,839 | 1 | 535 | 4,556 | 308 | 235 | 29 | 26,109 |
| 1998 - | - | 98 | 567 | 231 | $47^{3}$ | - | 28 | - | 26,273 | 13 | 131 | 5,278 | 228 | 211 | 94 | 33,199 |
| 1999 - | - | 108 | $61^{3}$ | 430 | 97 | 14 | 10 | - | 24,634 | 6 | 68 | 4,422 | 36 | 247 | 62 | 30,195 |
| 2000- | - | $67^{3}$ | 25 | 222 | 51 | 65 | 1 | - | 19,052 | 2 | 131 | 4,631 | 87 |  | $203{ }^{6}$ | 24,537 |
| 2001 - | - | $111{ }^{3}$ | 46 | 436 | 34 | 3 | 5 | - | 23,071 | 5 | 186 | 4,738 | 91 | Est | - $12239{ }^{6}$ | 28,965 |
| 2002 - | - | $135{ }^{3}$ | 89 | 141 | 49 | 44 | 4 | - | 10,713 | $8^{3}$ | 276 | 4,736 | $193{ }^{2}$ | 15 | $234{ }^{6}$ | 16,637 |
| 2003 Swed |  | $173{ }^{3}$ | 31 | 154 | $44^{3}$ | 9 | $5^{3}$ | 89 | 8,063 | 7 | 50 | 1,431 | $47^{2}$ | - | $258{ }^{6}$ | 10,361 |
| 20041 | - | 607 | $17^{3}$ | 78 | $24^{3}$ | 40 | 3 | 33 | 7,608 | 42 | 240 | 3,601 ${ }^{2}$ | $260^{2}$ |  | $146{ }^{6}$ | 12,699 |
| 2005 Can | Lith | 1,194 | 56 | 106 | $75^{3}$ | $12^{2}$ | $4^{3}$ | $55^{2}$ | 7,844 |  | 196 | 5,637 | $171{ }^{3}$ |  | $147{ }^{6}$ | 15,501 |
| 2006433 | 845 | 3,919 | 223 | 2,518 | $107^{3}$ | 2,544 | $4^{3} 12^{3}$ | 21 | 11,0152 | 2,496 | ²1,873 | 312,126 | $719{ }^{2}$ | 396 | 1,066 | 40,313 |
| 2007 Latv | 785 | 2,343 | 249 | 587 | $84^{3}$ | 1,647 |  | 20 | 8,9931 | 1,081 | 121,708 | 86,550 | 2,186 | $6^{2} 684$ | $257{ }^{6}$ | 27,181 |
| 2008267 | 117 | 2,123 ${ }^{3}$ | 250 | 46 | $74^{3}$ | $36^{3}$ | $2^{3}$ | 15 | 7,416 | 8 | 785 | 7,866 | 1,183 | $3^{2} \mathbf{E U}^{7}$ | $168{ }^{6}$ | 20,356 |
| 20091- | - | 1,413 | 19 | 100 | 72 | 76 | - | 4 | 8,149 | 338 | 836 | 4,541 | 177 | 889 | 113 | 16,727 |

${ }^{1}$ Provisional figures.
${ }^{2}$ Working Group figure.
${ }^{3}$ As reported to Norwegian authorities or NEAFC.
${ }^{4}$ Includes former GDR prior to 1991.
${ }^{5}$ USSR prior to 1991.
${ }^{6}$ UK(E\&W)+UK(Scot.)
${ }^{7} \mathrm{EU}$ not split on countries.

Table D2. REDFISH in Sub-area IV (North Sea). Nominal catch ( $\mathbf{t}$ ) by countries as officially reported to ICES. Not included in the assessment.

| Year | Belgium | Denmark | Faroe Islands | France | Germany | Ireland | Netherlands | Norway | Sweden | UK <br> (England \& Wales) | $\begin{array}{r} \mathrm{UK} \\ (\mathrm{Scotl}) \end{array}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | - | 24 | - | 578 | 183 | - | - | 1,048 | - | 35 | 1 | 1,869 |
| 1987 | - | 16 | 3 | 833 | 70 | - | - | 411 | - | 16 | 55 | 1,404 |
| 1988 | - | 32 | 90 | 915 | 188 | - | - | 696 | - | 125 | 9 | 2,055 |
| 1989 | 1 | 23 | 13 | 554 | 111 | - | - | $500{ }^{2}$ | - | 134 | 6 | 1,342 |
| 1990 | + | 41 | 25 | 554 | 47 | - | - | $483{ }^{2}$ | - | 369 | 6 | 1,525 |
| 1991 | 5 | 29 | 144 | 914 | 213 | - | 2 | $415^{2}$ | - | 43 | 38 | 1,803 |
| 1992 | 4 | 22 | 23 | 1,960 | 170 | - | 1 | 416 | - | 65 | 122 | 2,783 |
| 1993 | 28 | 14 | 4 | 1,211 | 33 | - | 1 | 373 | - | 138 | 71 | 1,873 |
| 1994 | 4 | 13 | 1 | 863 | 324 | - | 8 | 371 | - | 38 | 66 | 1,688 |
| 1995 | 16 | 12 | 65 | 1,120 | 80 | - | 16 | 297 | - | 46 | 241 | 1,893 |
| 1996 | 20 | 20 | 1 | 932 | 74 | - | 41 | 363 | - | 37 | 146 | 1,634 |
| 1997 | 16 | 23 | - | 1,049 | 45 | - | 53 | 595 | - | 21 | 528 | 2,330 |
| 1998 | 2 | 27 | 12 | 570 | 370 | 4 | 21 | 1,113 | - | 68 | 681 | 2,868 |
| 1999 | 3 | 52 | 1 | - | 58 | 39 | 16 | 862 | - | 67 | 465 | 1,563 |
| 2000 | 5 | 41 | - | 224 | 19 | 28 | 19 | 443 | - | 132 | 486 | 1,397 |
| 2001 | 4 | 96 | - | 272 | 13 | 19 | + | 421 | - | 80 | 458 | 1,363 |
| 2002 | 2 | 40 | 2 | 98 | 11 | 7 | + | 241 | - |  | $524{ }^{3}$ | 925 |
| 2003 | 1 | 71 | 2 | 26 | 2 | - | - | 474 | - | Portugal | $463{ }^{3}$ | 1,071 |
| 2004 | + | 42 | 3 | 26 | 1 | - | - | 287 | - | - | $214{ }^{3}$ | 578 |
| 2005 | 2 | 34 | - | 10 | 1 | - | - | 84 | - | - | $28^{3}$ | 159 |
| 2006 | 1 | 49 | 1 | 12 | 3 | - | - | 155 | - | 33 | $79^{3}$ | 333 |
| $2007{ }^{1}$ | + | 27 | - | 8 | 1 | - | - | 107 | + | - | $78{ }^{3}$ | 221 |
| $2008{ }^{1}$ | + | 3 | - | 35 | 1 | - | - | 77 | + | - | $54^{3}$ | 170 |
| $2009{ }^{1}$ | - | - | - | - | - | - | - | 120 | + | - | 87 | 207 |

[^5]Table D3. Sebastes mentella. Ave rage catch (numbers of specimens) per hour trawling of different ages of Sebastes mentella in the Russian groundfish survey in the Barents Sea and Svalbard areas (1976®1983 published in "Anna les Biologiques").

| Year class | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1974 | - | - | 4.8 | - | 4.9 | 22.8 | 4.8 | 4.8 | - | - | - | 3 |
| 1975 | - | 7.4 | - | 1.7 | 6.4 | 2.4 | 3.5 | 5 | - | - | 4 | - |
| 1976 | 7 | - | 8.1 | 1.2 | 2.5 | 6.8 | 4.9 | 5 | 1 | 13 | - | - |
| 1977 | - | 0.2 | 0.2 | 0.2 | 0.9 | 5.1 | 3.7 | 1 | 19 | 2 | - | - |
| 1978 | 0.8 | 0.02 | 0.9 | 1 | 5 | 3.8 | 2 | 20 | 6 | - | - | - |
| 1979 | - | 1.9 | 1.4 | 3.6 | 2.3 | 9 | 11 | 16 | 1 | - | - | 0.1 |
| 1980 | 0.3 | 0.4 | 2 | 2.5 | 16 | 6 | 11 | 25 | 2 | - | 1.5 | 2 |
| 1981 | - | 2.2 | 3.9 | 20 | 6 | 12 | 47 | 18 | 6.3 | 1.6 | 0.5 | 1 |
| 1982 | 198 | 132 | 13 | 15 | 34 | 44 | 39 | 32.6 | 4.3 | 3.1 | 4.9 | + |
| 1983 | 125 | 3 | 5 | 6 | 31 | 34 | 323 | 133 | 4 | 4.2 | 0.6 | 1.1 |
| 1984 | - | 10 | 2 | - | 5 | 183 | 19 | 2.2 | 2.4 | 0.2 | 1.7 | 2.4 |
| 1985 | 107 | 7 | - | 1 | 5.2 | 162 | 1.7 | 1.7 | 0.6 | 2.8 | 3.8 | 0.3 |
| 1986 | 2 | - | 1 | 1.8 | 8.4 | 3.6 | 2.1 | 1.2 | 5.6 | 8.2 | 0.9 | 0.7 |
| 1987 | - | 3 | 379 | 1.3 | 8 | 4.1 | 2 | 10.6 | 9.6 | 1.4 | 2 | 1.3 |
| 1988 | 4 | 58.1 | 4.3 | 133 | 258 | 3.9 | 8.6 | 112 | 2.8 | 4.2 | 3 | 4.7 |
| 1989 | 8.7 | 9 | 17 | 23.4 | 4.6 | 5.4 | 4 | 6.6 | 6.6 | 4.1 | 7.7 | 5.3 |
| 1990 | 2.5 | 6.3 | 6.1 | 1 | 4.3 | 1.7 | 115 | 6.5 | 5.5 | 6.7 | 7.4 | 3.6 |
| 1991 | 0.3 | 1 | 0.5 | 1.5 | 1.2 | 113 | 3.9 | 3.3 | 4.6 | 5.8 | 2.7 | 1.9 |
| 1992 | 0.6 | + | 0.2 | 0.1 | 4.3 | 1.3 | 2 | 2.3 | 4.9 | 2.3 | 1 | 4.1 |
| $1993{ }^{1}$ | - | + | 1.5 | 1.8 | 1 | 1.2 | 3 | 4.2 | 2.6 | 2 | 3.2 | 2.1 |
| 1994 | 0.3 | 3.5 | 1.7 | 1.7 | 0.9 | 3.6 | 5.2 | 4.3 | 3.1 | 3.3 | 1.8 | 1.2 |
| 1995 | 2.8 | 1 | 1.1 | 0.4 | 2.2 | 2.6 | 3.5 | 3.4 | 2.9 | 1.2 | 1 | 8.5 |
| $1996{ }^{2}$ | + | 0.1 | 0.1 | 0.4 | 0.7 | 1.1 | 1 | 1.4 | 1 | 0.8 | 3.7 | 0.6 |
| 1997 | - | - | + | 0.4 | 0.5 | 0.3 | 0.9 | 0.6 | 1 | 1.1 | 0.5 | 0.4 |
| 1998 | - | 0.1 | 0.2 | 0.3 | 0.2 | 1.1 | 0.5 | 0.7 | 1 | 0.4 | 0.4 | 0.7 |
| 1999 | 0.1 | - | 0.1 | + | 0.1 | 0.3 | 0.5 | 0.8 | 0.5 | 0.2 | 0.4 |  |
| 2000 | - | 0.6 | 0.1 | 0.5 | 0.3 | 0.3 | 0.6 | 0.4 | 0.1 | 0.1 |  |  |
| 2001 | - | 0.1 | 0.4 | - | 0.1 | 0.2 | 0.2 | 0.3 | 0.2 |  |  |  |
| $2002^{3}$ | 0.1 | 0.5 | 0.1 | - | - | 0.1 | 0.5 | 0.4 |  |  |  |  |
| 2003 | - | - | 0.1 | - | 0.3 | 1.0 | 0.5 |  |  |  |  |  |
| 2004 | - | 0.2 | 0.3 | 0.5 | 1.5 | 0.9 |  |  |  |  |  |  |
| 2005 | - | - | 1.4 | 1.9 | 1.4 |  |  |  |  |  |  |  |
| $2006{ }^{4}$ | 0.1 | 1.8 | 1.2 | 1.1 |  |  |  |  |  |  |  |  |
| 2007 | 2.5 | 0.4 | 0.1 |  |  |  |  |  |  |  |  |  |
| 2008 | 0.1 | 0.1 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1.6 |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{1}$ - Not com <br> ${ }^{2}$ - Area sur <br> ${ }^{3}$ - Area sur <br> ${ }^{4}$ - Area surv | te are ed res ed res ed rest | verag <br> ted to <br> ted to <br> ted to | Divis <br> barea <br> barea <br> ision | IIb. <br> d Divi <br> d Divi <br> and IIb | IIa o <br> IIb on <br> ly. |  |  |  |  |  |  |  |

Table D4a. Sebastes mentella ${ }^{1}$ in Division IIb. Abundance indices (on length) from the bottom trawl survey in the Svalbard area (Division IIb) in summer/fall 1986-2009 (numbers in millions).

| Length group (cm) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 5.0-9.9 | $\begin{array}{r} 10.0- \\ 14.9 \end{array}$ | $\begin{array}{r} 15.0- \\ 19.9 \end{array}$ | $\begin{gathered} 20.0- \\ 24.9 \end{gathered}$ | $\begin{array}{r} 25.0- \\ 29.9 \end{array}$ | $\begin{array}{r} 30.0- \\ 34.9 \end{array}$ | $\begin{gathered} 35.0- \\ 39.9 \end{gathered}$ | $\begin{gathered} 40.0- \\ 44.9 \end{gathered}$ | >45.0 | Total |
| $1986{ }^{2}$ | 6 | 101 | 192 | 17 | 10 | 5 | 2 | 4 | + | 338 |
| $1987{ }^{2}$ | 20 | 14 | 140 | 19 | 6 | 2 | 1 | 2 | + | 208 |
| $1988{ }^{2}$ | 33 | 23 | 82 | 77 | 7 | 3 | 2 | 2 | + | 228 |
| 1989 | 566 | 225 | 24 | 72 | 17 | 2 | 2 | 8 | 4 | 921 |
| 1990 | 184 | 820 | 59 | 65 | 111 | 23 | 15 | 7 | 3 | 1,287 |
| 1991 | 1,533 | 1,426 | 563 | 55 | 138 | 38 | 30 | 7 | 1 | 3,791 |
| 1992 | 149 | 446 | 268 | 43 | 22 | 15 | 4 | 7 | 4 | 958 |
| 1993 | 9 | 320 | 272 | 89 | 16 | 13 | 3 | 1 | + | 722 |
| 1994 | 4 | 284 | 613 | 242 | 10 | 9 | 2 | 2 | 1 | 1,165 |
| 1995 | 33 | 33 | 417 | 349 | 77 | 18 | 5 | 1 | + | 933 |
| 1996 | 56 | 69 | 139 | 310 | 97 | 8 | 4 | 1 | 1 | 685 |
| 1997 | 3 | 44 | 13 | 65 | 57 | 9 | 5 | + | + | 195 |
| 1998 | + | 37 | 35 | 28 | 132 | 73 | 45 | 2 | + | 353 |
| 1999 | 4 | 3 | 121 | 62 | 259 | 169 | 42 | 1 | 0 | 661 |
| 2000 | + | 10 | 31 | 59 | 126 | 143 | 21 | 1 | 0 | 391 |
| 2001 | 1 | 5 | 3 | 32 | 57 | 228 | 50 | 3 | 0 | 378 |
| 2002 | 1 | 4 | 6 | 21 | 62 | 266 | 47 | 4 | + | 410 |
| 2003 | 1 | 5 | 7 | 11 | 56 | 271 | 50 | 1 | 0 | 403 |
| 2004 | 0 | 2 | 7 | 6 | 14 | 78 | 53 | 2 | 0 | 163 |
| 2005 | 1 | 1 | 6 | 11 | 19 | 93 | 63 | 1 | 0 | 196 |
| 2006 | 82 | 6 | 5 | 7 | 49 | 211 | 101 | 3 | 0 | 463 |
| 2007 | 98 | 68 | 1 | 5 | 11 | 95 | 109 | 3 | 0 | 387 |
| 2008 | 119 | 45 | 20 | 3 | 9 | 25 | 79 | 4 | 0 | 303 |
| 2009 | 53 | 305 | 228 | 34 | 9 | 63 | 328 | 9 | 0 | 1029 |

${ }^{1}$ - Includes some unidentified Sebastes specimens, mostly less than $\mathbf{1 5} \mathbf{~ c m}$.
${ }^{2}$ - Old trawl equipment (bobbins gear and 80 meter sweep length)

Table D4b. Sebastes mentella ${ }^{1}$ in Division IIb. Norwegian bottom trawl survey indices (on age) in the Svalbard area (Division IIb) in summer/fall 1992-2009 (numbers in millions).

| Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | Total |
| 1992 | 283 | 419 | 484 | 131 | 58 | 45 | 14 | 8 | 5 | 2 | 7 | 2 | 1 | 3 | 1,462 |
| 1993 | 2 | 527 | 117 | 202 | 142 | 8 | 23 | 6 | 13 | 1 | 7 | 1 | 1 | + | 1,050 |
| 1994 | 7 | 280 | 290 | 202 | 235 | 42 | 94 | 1 | 1 | 3 | 4 | 1 | 1 | + | 1,161 |
| 1995 | 4 | 50 | 365 | 237 | 132 | 61 | 19 | 17 | 11 | + | 1 | 3 | 0 | 0 | 900 |
| 1996 | 23 | 47 | 15 | 37 | 105 | 144 | 84 | 17 | 51 | 32 | 34 | 9 | 6 | 2 | 605 |
| 1997 | 8 | 43 | 6 | 6 | 40 | 20 | 30 | 25 | 7 | 3 | 1 | 2 | 2 | 1 | 194 |
| 1998 | + | 26 | 28 | 14 | 10 | 13 | 69 | 66 | 49 | 15 | 1 | 6 | 15 | 5 | 317 |
| 1999 | 3 | 16 | 114 | 27 | 36 | 53 | 117 | 78 | 67 | 41 | 45 | 11 | 19 | 13 | 640 |
| 2000 | 4 | 6 | 6 | 14 | 35 | 22 | 31 | 54 | 81 | 60 | 24 | 24 | 10 | 8 | 379 |
| 2001 | 2 | 4 | 3 | 1 | 9 | 16 | 22 | 30 | 34 | 57 | 57 | 50 | 54 | 6 | 344 |
| 2002 | 3 | 2 | 4 | 2 | 5 | 22 | 34 | 23 | 88 | 36 | 62 | 64 | 15 | 21 | 379 |
| 2003 | 0.3 | 3 | 4 | 3 | 5 | 4 | 29 | 31 | 50 | 59 | 45 | 70 | 38 | 23 | 365 |
| 2004 | 1 | 1 | 3 | 3 | 1 | 4 | 2 | 9 | 9 | 18 | 15 | 17 | 19 | 9 | 113 |
| 2005 | 1 | 1 | 2 | 3 | 3 | 6 | 9 | 15 | 14 | 16 | 14 | 21 | 22 | 25 | 152 |
| 2006 | 33 | 1 | 3 | 3 | 2 | 9 | 17 | 27 | 24 | 35 | 29 | 45 | 25 | 34 | 287 |
| 2007 | 23 | 45 | 0 | 0 | 3 | 2 | 5 | 5 | 8 | 5 | 5 | 9 | 29 | 19 | 158 |
| 2008 | 6 | 22 | 22 | 12 | 1 | 2 | 2 | 5 | 4 | 4 | 3 | 5 | 10 | 6 | 102 |
| 2009 | 82 | 132 | 146 | 73 | 92 | 50 | 32 | 2 | 5 | 6 | 3 | 2 | 36 | 92 | 752 |

${ }^{1}$ - Includes some unidentified Sebastes specimens, mostly less than 15 cm .

Table D5a. Sebastes mentella ${ }^{1}$. Abundance indices (on length) from the bottom trawl surve ys in the Barents Sea in the winter 1986-2010 (numbers in millions). The area coverage was extended from 1993 onwards.

|  |  |  |  |  |  | Length group (cm) |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | $5.0-9.9$ | $10.0-$ | $15.0-$ | $20.0-$ | $25.0-$ | $30.0-$ | $35.0-$ | $40.0-$ | $>45.0$ | Total |
| 1986 | 81 | 152 | 205 | 88 | 169 | 130 | 88 | 24 | 13.8 | 950 |
| 1987 | 72 | 25 | 227 | 56 | 35 | 11 | 5 | 1 | 0.1 | 433 |
| 1988 | 587 | 25 | 133 | 182 | 40 | 50 | 48 | 4 | 0.1 | 1068 |
| 1989 | 623 | 55 | 28 | 177 | 58 | 9 | 8 | 2 | 0.3 | 961 |
| 1990 | 324 | 305 | 36 | 56 | 80 | 13 | 13 | 2 | 0.2 | 828 |
| 1991 | 395 | 449 | 86 | 39 | 96 | 35 | 24 | 3 | 0.2 | 1127 |
| 1992 | 139 | 367 | 227 | 35 | 55 | 34 | 8 | 2 | 0.5 | 867 |
| 1993 | 31 | 593 | 320 | 116 | 24 | 25 | 6 | 1 | + | 1117 |
| 1994 | 7 | 259 | 289 | 284 | 51 | 70 | 20 | 1 | 0.1 | 982 |
| 1995 | 264 | 71 | 638 | 506 | 91 | 69 | 31 | 4 | 0.5 | 1674 |
| 1996 | 213 | 100 | 191 | 338 | 134 | 42 | 17 | 1 | 0.3 | 1037 |
| $1997^{2}$ | 63 | 121 | 25 | 278 | 274 | 72 | 41 | 5 | 0.2 | 879 |
| $1998^{2}$ | 1 | 91 | 63 | 101 | 203 | 41 | 13 | 2 | 0.2 | 514 |
| 1999 | 2 | 7 | 68 | 37 | 167 | 72 | 21 | 3 | 0.1 | 377 |
| 2000 | 9 | 13 | 39 | 77 | 142 | 97 | 27 | 7 | 1.5 | 412 |
| 2001 | 9 | 22 | 7 | 55 | 77 | 73 | 9 | 1 | 0.1 | 254 |
| 2002 | 16 | 7 | 19 | 42 | 104 | 114 | 23 | 1 | + | 326 |
| 2003 | 4 | 4 | 10 | 13 | 71 | 200 | 47 | 6 | 0.3 | 354 |
| 2004 | 2 | 3 | 7 | 19 | 33 | 87 | 32 | 2 | 0.1 | 184 |
| 2005 | + | 6 | 7 | 11 | 28 | 153 | 87 | 4 | 0.2 | 297 |
| 2006 | 99 | 2 | 10 | 15 | 23 | 103 | 82 | 3 | 0.7 | 336 |
| 2007 | 446 | 125 | 3 | 6 | 12 | 119 | 120 | 7 | 0.2 | 838 |
| 2008 | 846 | 354 | 26 | 5 | 12 | 114 | 180 | 5 | 0.1 | 1542 |
| 2009 | 34 | 101 | 70 | 2 | 5 | 26 | 65 | 2 | 0.1 | 304 |
| 2010 | 647 | 273 | 213 | 64 | 7 | 73 | 190 | 6 | 0.4 | 1474 |

[^6]Table D5b. Sebastes mentella ${ }^{1}$ in Sub-areas I and II. Preliminary Norwegian bottom trawl indices (on age) from the annual Barents Sea survey in February 1992-2009 (numbers in millions). The area coverage was extended from 1993 onwards.

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | Total |
| 1992 | 351 | 252 | 132 | 56 | 14 | 11 | 3 | 9 | 18 | 16 | 12 | 11 | 2 | 5 | 892 |
| 1993 | 38 | 473 | 192 | 242 | 62 | 45 | 19 | 22 | 13 | 11 | 10 | 4 | 2 | 3 | 1,136 |
| 1994 | 7 | 85 | 332 | 189 | 370 | 228 | 73 | 42 | 3 | 30 | 8 | 14 | 25 | 7 | 1,413 |
| 1995 | 308 | 45 | 146 | 264 | 364 | 211 | 69 | 23 | 7 | 17 | 23 | 9 | 11 | 10 | 1,507 |
| 1996 | 173 | 119 | 109 | 114 | 128 | 122 | 106 | 64 | 24 | 19 | 12 | 7 | 8 | 4 | 1,009 |
| $1997^{2}$ | 43 | 101 | 19 | 54 | 96 | 43 | 44 | 171 | 76 | 74 | 39 | 29 | 10 | 9 | 808 |
| $1998^{2}$ | 1 | 73 | 49 | 27 | 13 | 52 | 107 | 104 | 41 | 18 | 7 | 4 | 3 | 3 | 502 |
| 1999 | 1 | + | 32 | 43 | 30 | 24 | 30 | 81 | 79 | 28 | 2 | 1 | 6 | + | 357 |
| 2000 | 9 | 12 | 21 | 17 | 9 | 39 | 77 | 73 | 50 | 41 | 14 | 10 | 7 | 6 | 385 |
| 2001 | 1 | 17 | 8 | 1 | 7 | 22 | 39 | 30 | 34 | 23 | 24 | 17 | 9 | 3 | 236 |
| 2002 | 18 | 4 | 12 | 7 | 4 | 14 | 49 | 55 | 27 | 19 | 34 | 24 | 28 | 11 | 306 |
| 2003 | 0 | 2 | 2 | 4 | 6 | 6 | 14 | 39 | 24 | 34 | 39 | 65 | 46 | 20 | 301 |
| 2004 | 0 | 2 | 3 | 1 | 9 | 12 | 15 | 20 | 36 | 8 | 28 | 3 | 25 | 12 | 172 |
| 2005 | 0 | 4 | 3 | 3 | 6 | 6 | 11 | 15 | 23 | 14 | 21 | 40 | 35 | 49 | 229 |
| 2006 | 4 | 1 | 5 | 5 | 5 | 8 | 15 | 12 | 6 | 15 | 21 | 17 | 32 | 36 | 180 |
| 2007 | 428 | 82 | 13 | 1 | 2 | 2 | 5 | 7 | 8 | 8 | 21 | 20 | 31 | 35 | 144 |
| 2008 | 648 | 173 | 107 | 11 | 0 | 2 | 5 | 7 | 5 | 10 | 10 | 28 | 27 | 40 | 1073 |
| 2009 | 39 | 37 | 32 | 28 | 30 | 14 | 2 | 4 | 1 | 2 | 3 | 3 | 3 | 3 | 203 |

${ }^{1}$ - Includes some unidentified Sebastes specimens, mostly less than 15 cm .
${ }^{2}$ - Adjusted indices to account for not covering the Russian EEZ in Subarea I.

Table D6. Sebastes men tella in Sub-areas I and II. Abundance indices (on age) from the Ecosystem survey in August-September 1996-2009 covering the Norwegian Economic Zone (NEZ) and Svalbard incl. the area north and east of Spitsbergen (numbers in thousands and total biomass in thousand tonnes) and the continental slope down to 1500 m .

| Year | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ | Total N | Total B |
| 1996 | 146198 | 112742 | 22353 | 53507 | 165531 | 181980 | 108738 | 43328 | 65310 | 40546 | 38254 | 19843 | 29446 | 10931 | 17414 | 1056120 | 171 |
| 1997 | 62682 | 130816 | 12492 | 23452 | 74342 | 55880 | 76607 | 82503 | 17640 | 14274 | 675 | 2238 | 1723 | 633 | 8765 | 564723 | 73 |
| 1998 | 313 | 78767 | 85715 | 39849 | 25805 | 23413 | 84825 | 100332 | 54287 | 24329 | 11334 | 7457 | 15250 | 576 | 25212 | 577464 | 105 |
| 1999 | 5359 | 23240 | 117170 | 47851 | 41608 | 76797 | 128677 | 73306 | 58018 | 64781 | 49890 | 13565 | 18458 | 12171 | 24672 | 755562 | 155 |
| 2000 | 5964 | 23169 | 14336 | 19960 | 52666 | 68081 | 83857 | 77513 | 100442 | 72294 | 71148 | 36599 | 17183 | 20590 | 26501 | 690837 | 178 |
| 2001 | 5026 | 6541 | 10957 | 1093 | 19766 | 25591 | 36594 | 51644 | 44407 | 61704 | 50083 | 86122 | 53952 | 15699 | 31877 | 507131 | 162 |
| 2002 | 9112 | 6646 | 7379 | 3821 | 8635 | 28215 | 47456 | 63903 | 103368 | 49964 | 76133 | 71970 | 25241 | 36765 | 34957 | 573565 | 181 |
| 2003 | 3954 | 7394 | 6142 | 3540 | 8030 | 9388 | 48564 | 59051 | 98554 | 69901 | 83192 | 73521 | 69970 | 37162 | 47323 | 625687 | 213 |
| 2004 | 9068 | 10837 | 9008 | 7292 | 2510 | 7896 | 8193 | 15268 | 25544 | 29654 | 35249 | 21142 | 39581 | 25976 | 66792 | 314030 | 111 |
| 2005 | 1310 | 4406 | 5241 | 5031 | 5722 | 8740 | 13452 | 20672 | 16207 | 19353 | 17430 | 32028 | 37564 | 34815 | 57103 | 279072 | 103 |
| 2006 | 156578 | 5162 | 6695 | 5217 | 3768 | 10754 | 18771 | 29174 | 25278 | 38958 | 31869 | 46885 | 30895 | 44299 | 147951 | 602255 | 184 |
| 2007 | 302988 | 224153 | 290 | 7686 | 11346 | 2031 | 7903 | 10770 | 12182 | 6578 | 6367 | 9998 | 41425 | 22090 | 211178 | 876986 | 172 |
| 2008 | 86880 | 183796 | 121430 | 21430 | 4178 | 3009 | 3334 | 6991 | 5120 | 4441 | 3581 | 6008 | 10352 | 10172 | 99808 | 570530 | 89 |
| $2009{ }^{1}$ | 68775 | 161824 | 159311 | 89730 | 109995 | 46726 | 21792 | 2560 | 15734 | 23819 | 8737 | 11244 | 11739 | 42990 | 213123 | 999245 | 200 |

${ }^{1}$ in 2009, data in the arctic waters are not included

| Year | Period of survey | Age ${ }^{\text {atal }}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Area of survey <br> in <br> n.m. ${ }^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\begin{aligned} & 1- \\ & 4 \end{aligned}$ |  | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21+ | $\begin{aligned} & \text { Numbe } \\ & \text { rs } \\ & 10^{6} \end{aligned}$ | Biomas s t $10^{3}$ | $\begin{aligned} & \hline \text { SSN } \\ & 10^{6} \end{aligned}$ | $\begin{aligned} & \hline \text { SSB } \\ & \mathrm{t} \\ & 10^{3} \end{aligned}$ |  |
| 1992 | April | 29 | 27 | 27 | 37 | 36 | 50 | 78 | 39 | 34 | 40 | 44 | 43 | 28 | 17 | 13 | 4 | 7 | 3 | 566 | 218 | 191 | 114 | 25300 |
| 1993 | April | 31 | 15 | 13 | 6 | 6 | 20 | 56 | 56 | 38 | 28 | 29 | 27 | 19 | 12 | 7 | 3 | 1 | 2 | 396 | 150 | 151 | 90 | 23500 |
| 1994 | No Data |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1995 | May | + | 32 | 51 | 83 | 90 | 41 | 31 | 31 | 41 | 94 | 73 | 48 | 30 | 10 | 9 | 4 | 1 | + | 669 | 202 | 211 | 102 | 23300 |
| 1996 | No Data |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1997 | Apr-May | 86 | 6 | 24 | 102 | 150 | 53 | 48 | 24 | 20 | 26 | 36 | 28 | 11 | 9 | 4 | 2 | 1 | + | 630 | 170 | 111 | 58 | 22400 |
| 1998 | April | 1 | + | 8 | 47 | 77 | 63 | 71 | 46 | 27 | 19 | 23 | 23 | 25 | 6 | 3 | 2 | 1 | + | 442 | 153 | 106 | 57 | 22931 |
| 1999 | Apr-May | 11 | 1 | 9 | 14 | 57 | 75 | 63 | 73 | 31 | 25 | 17 | 15 | 11 | 8 | 3 | 1 | 1 | 1 | 415 | 134 | 120 | 55 | 19333 |
| 2000 | Apr-May | 2 | 2 | 14 | 15 | 62 | 100 | 143 | 122 | 54 | 34 | 24 | 29 | 12 | 11 | 7 | 2 | 1 | 1 | 635 | 208 | 114 | 53 | 22000 |
| 2001 | Apr-May | 11 | 1 | 11 | 22 | 24 | 84 | 123 | 134 | 144 | 115 | 78 | 40 | 27 | 19 | 10 | 4 | + | 3 | 850 | 316 | 339 | 152 | 23000 |
| 2002 | No Data |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2003 | No Data |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2004 | No Data |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2005 | No Data |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2006 | No Data |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2007 | No Data |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2008 | No Data |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2009 | No Data |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table D8a. Sebastes mentella in Sub-areas I and II. Maturity ogives from Russian research vessels. Sexes combined. Data collected during April-June in the Kopytov area (western Barents Sea) and adjacent waters.

| AGE | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | 1993 | 1995 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.018 | 0.021 | 0.000 | 0.000 | 0.000 |
| 8 | 0.000 | 0.000 | 0.000 | 0.046 | 0.000 | 0.000 | 0.000 | 0.000 | 0.014 | 0.016 | 0.000 | 0.000 |
| 9 | 0.000 | 0.000 | 0.012 | 0.139 | 0.013 | 0.033 | 0.000 | 0.027 | 0.000 | 0.059 | 0.048 | 0.082 |
| 10 | 0.028 | 0.074 | 0.131 | 0.174 | 0.092 | 0.133 | 0.055 | 0.130 | 0.074 | 0.110 | 0.087 | 0.196 |
| 11 | 0.125 | 0.178 | 0.300 | 0.138 | 0.169 | 0.364 | 0.111 | 0.312 | 0.171 | 0.333 | 0.202 | 0.405 |
| 12 | 0.297 | 0.473 | 0.688 | 0.358 | 0.396 | 0.480 | 0.368 | 0.281 | 0.276 | 0.579 | 0.375 | 0.442 |
| 13 | 0.562 | 0.684 | 0.714 | 0.470 | 0.452 | 0.696 | 0.587 | 0.566 | 0.622 | 0.689 | 0.489 | 0.442 |
| 14 | 0.760 | 0.716 | 0.824 | 0.637 | 0.761 | 0.925 | 0.696 | 0.736 | 0.714 | 0.788 | 0.742 | 0.648 |
| 15 | 0.855 | 0.794 | 0.848 | 0.762 | 0.939 | 0.962 | 0.729 | 0.831 | 0.871 | 0.813 | 0.833 | 0.775 |
| 16 | 1.000 | 1.000 | 1.000 | 1.000 | 0.886 | 0.953 | 0.789 | 0.958 | 0.919 | 0.903 | 0.904 | 0.865 |
| 17 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 0.977 | 1.000 | 0.950 | 1.000 | 0.923 | 1.000 | 0.909 |
| 18 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |

Table D8b. Sebastes mentella in Sub-areas I and II. Modelled maturit y ogive from Norwegian research vessels. Sexes combined. Data collected during the pelagic summer surveys (2007/2008), the slope surve y (October 2008), the Norwegian part of the ecosystem surveys in the Barents Sea (summer 2004-2008) and the winter surveys (2004-2008).

| AGE | MATURITY |
| :---: | :--- |
| 1 | 0.00 |
| 2 | 0.00 |
| 3 | 0.00 |
| 4 | 0.00 |
| 5 | 0.00 |
| 6 | 0.01 |
| 7 | 0.03 |
| 8 | 0.06 |
| 9 | 0.13 |
| 10 | 0.27 |
| 11 | 0.48 |
| 12 | 0.60 |
| 13 | 0.71 |
| 14 | 0.79 |
| 15 | 0.86 |
| 16 | 0.91 |
| 17 | 0.94 |
| 18 | 0.96 |
| 19 | 0.98 |
| 20 | 0.98 |
| 21 | 0.99 |
| 22 | 0.99 |
| $23+$ | 1.00 |
|  |  |



Figure D1. Map showing the strata system, the specific pelagic 0-group trawl stations and the abundance of 0 -group Sebastes mentella during the joint Norwegian-Russian Ecosystem survey in the Barents Sea and Svalbard. Example from 2008.


Figure D2. Overview of different scientific surveys contributing with information about the Sebastes mentella stock in Sub-area I and II in 2008-2009. A: Norwegian-Russian survey in winter 2008, B: Norwegian slope survey in March/April 2009, C: Norwegian-Russian ecosystem survey in summer 2008, D: Norwegian slope survey in August 2008, E: Norwegian slope survey in November 2008, F: Russian survey in October/December 2008.

## 7 Golden redfish (Sebastes marinus) in Subareas I and II

### 7.1 Status of the Fisheries

### 7.1.1 Recent regulations of the fishery

A description of the historical development of the fishery and regulations is found in the Quality handbook for this stock. The Handbook has been updated in 2010 (see Annex in this report).

Prior to 1 January 2003 there were no regulations particularly for the S. marinus fishery, and the regulations aimed at S. mentella (see chapter 6.1.1) had only marginal effects on the $S$. marinus stock. After this date, all directed trawl fishery for redfish (both S. marinus and S. mentella) outside the permanently closed areas have been forbidden in the Norwegian Economic Zone north of $62^{\circ} \mathrm{N}$ and in the Svalbard area. When fishing for other species it is currently legal to have up to $15 \%$ redfish (both species together) in round weight as by catch per haul and on board at any time. Until 14 April 2004 there were no regulations of the other gears/fleets fishing for S. marinus. After this date, a minimum legal catch size of 32 cm has been set for all fisheries, with the allowance to have up to $10 \%$ undersized (i.e., less than 32 cm ) specimens of S. marinus (in numbers) per haul. In addition, a limited moratorium has been enforced in the conventional fisheries (gillnet, longline, handline, Danish seine) except for handline vessels less than 11 meters. Since 2007 this moratorium has been during 5 months, i.e., March-June and September, a change from April-May and September in 2006, 20 April-19 June in 2005 and 1-31 May in 2004. When fishing for other species (also during the moratorium) it is allowed to have up to $15 \%$ bycatch of redfish (in round weight) summarized during a week fishery from Monday to Sunday.

### 7.1.2 Landings prior to 2010 (Tables 7.1-7.4, D1 \& D2, Figures 7.1-7.2)

Nominal catches of S. marinus by country for Sub-areas I and II combined, and for each Sub-area and Division are presented in Tables 7.1-7.4. The total landings for both S. marinus and S. mentella are presented in Tables D1 and D2. Landings of S. marinus showed a decrease from a level of 23,000-30,000 t in 1984-1990 to a stable level of about $16,000-19,000 t$ in the years 1991-1999. Since then the landings have decreased further, and the total landings figures for S. marinus in 2003-2007 have been low but remarkable stable between $7,000-7,800 \mathrm{t}$. Provisional figures for 2009 indicate a further decline in landings down to 6,293 $t$, the lowest since the mid-1940ies. This is mainly attributable to decreasing catches in Division IIa (minus 226 t) and Sub-area I (minus 41 t). No significant changes in landings can be observed in Division Ilb. The time series of S. marinus landings is given in Figure 7.1 and shows a long-term (1908-2009) mean of $16,734 \mathrm{t}$.

The Norwegian landings are presented by gear and month in Figures 7.2a,b. Reported landings have diminished in 2009 for trawl and Danish seine and increased in gillnet (for the second year in a row) and longline. Since 2003, the limited moratorium for conventional gears seems to have reduced the catches taken by these gears from about $5,900 t$ to about $3,200 \mathrm{t}$ in 2007, but this trend has halted due to the increase in gillnet catches in 2008 and 2009. Due to the increase in catches by gillnets from 2,649 t in 2008 to $2,841 \mathrm{t}$ in 2009, the total catches, except trawl, have increased to $4,135 \mathrm{t}$. For fishing gears other than gillnet, bycatches in 2008 and 2009 are the lowest observed for the period 2003-2009.

The reported Russian catches of S. marinus were decreasing from 890 t in 2007 and 749 t in 2008 to 698 t in 2009 (Table 7.1)

The bycatch estimates of redfish (Sebastes spp.) in the Norwegian Barents Sea shrimp fisheries during 1983-2002 are completely dominated by S. mentella, and hence will influence the $S$. marinus to a much lesser extent. However, it probably inflicted an extra mortality on S. marinus in the coastal areas before the sorting grid was enforced in 1990. From 1 January 2006, the maximum authorised bycatch of redfish juveniles in the international shrimp fisheries in the northeast Arctic has been reduced from ten to three redfish per 10 kg shrimp.
Information describing the splitting of the redfish landings by species and area is given in the Quality handbook.

### 7.1.3 Expected landings in 2010

In 2009, total Norwegian catch ( $5,383 \mathrm{t}$, provisional figure) and total Russian catch ( 698 t ) are close to the values expected in the previous year. Under similar assumptions (reports from the first months of the year, a legal by-catch of $15 \%$ in all trawl fisheries, and an assumed effect of the regulations for the other gears) the Norwegian and Russian landings in 2010 are expected to be similar to those reported in 2009.

### 7.2 Data Used in the Assessment (Figure D 1)

An overview of the sampling levels (by season, area and gear) of the data used in the assessment is presented in Figure Dl for 2008. Of technical reasons it became impossible to present the same figure for 2009 before the report had to be printed. The sampling of S. marinus commercial catches should be improved. In 2009, only 3 of 11 metiers (area-quarter-gear combinations) responsible for $50 \%$ of the Norwegian landings were properly covered with age samples, compared to 8 of 13 metiers the year before

### 7.2.1 Catch-per-unit-effort (Table D11, Figure 7.3)

The CPUE-series for S. marinus from Norwegian 32-50 meter freezer trawlers and Factory trawlers ( $>53 \mathrm{~m}$ ) is presented from 1992 onwards (Table D11, Figure 7.3). Only data from days with more than $10 \%$ S. marinus in the catches (in weight) are included in the annual averages. Mean CPUEs with standard errors together with number of vessel days meeting the $10 \%$ criterion are presented in Table D11. Provisional figures for 20062009 indicate an important reduction in the effort of freezer trawlers since 2006 in comparison with the previous decade. The effort of factory trawlers has remained stable around 150 days since 2003.

Although the trawl fishery until 2003 was almost unregulated, the trawlers experienced fewer and fewer fishing days with more than $10 \%$ of their catches composed of S. marinus (Figure 7.3). During 2001-2005 both the catch-rates and the number of vessel-days were rapidly decreasing, and this is worrying since the criterion for defining it to be a S. marinus vessel-day have not been more than $20 \%$ (since 2003) or $10 \%$ (since 2004) S. marinus in each trawl haul. Since 2005 a slight improvement of the catch-rates is seen for both trawler fleets, but it is worrying that the number of vessel days containing a minimum of $10 \%$ redfish still are decreasing in one of the fleets. With some variation, the average annual catch-rates for the freezer trawlers have decreased from an average level of $350 \mathrm{~kg} /$ /rawl hour during mid 1990 ies to about $150 \mathrm{~kg} / \mathrm{h}$ since 2003, i.e., less than $40 \%$ of the former recent level.

Corresponding values for the factory trawlers are $600 \mathrm{~kg} / \mathrm{trawl}$ until 2001 and about $200-300 \mathrm{~kg} / \mathrm{h}$ since 2002. The decrease seems though to have halted for both fleets.

### 7.2.2 Catch at length and age (Table 7.5)

Catch at age data for 2006-2008 were revised. Age composition data for 2009 were only provided by Norway, accounting for $86 \%$ of the total landings. Other countries were assumed to have the same relative age distribution and mean weight as Norway. The updated catch-in-numbers at age matrix is shown in Table 7.5. Catch at length data were available from Norway and Portugal.

### 7.2.3 Weight at Age (Table 7.6)

Weight-at-age data for ages $7-24+$ were available from the Norwegian landings in 2009. Variations in the weight-at-age of young individuals $(<10 y)$ must be considered with caution as these numbers are derived from only a small number of aged individuals.

### 7.2.4 Maturity at age (Figure 7.7)

A maturity ogive has previously not been available for $S$. marinus, and knife-edge maturity at age 15 (age 15 as $100 \%$ mature) has hence been assumed. The improved maturity ogive modelled by the Gadget model, and based on maturation data (by length and age) collected from Norw egian surveys and landings, is presented (Figure 7.7). This analysis shows that $50 \%$ of the fish are mature at age 12. In previous years the maturity ogive was stable from the mid 1990s, however it was less reliable early in the modelled period. This was due to the maturity data the model was tuned to beginning in 1993. Large immature fish in the model before this would become mature before the data series started, and thus incur no penalty during optimisation. As a result the model over-predicted large immature fish in the early part of the time series, and under-predicted large mature fish for the same period. To rectify this, the maturity at age data for 1993-1995 w as averaged and input as "data" between 1986 and 1992. This was found to produce consistent maturity ogives in the model, as shown in Figure 7.7. Testing showed that this did not otherwise alter the model dynamics (note that no SSB-recruitment relationship is used in the model), and has therefore been adopted from the 2009 WG onwards.

### 7.2.5 Survey results (Tables D12a,b-D13a,b-D14, Figures 7.4a,b-7.5a,b)

The results from the following research vessel survey series were evaluated by the Working Group:

1 ) Norwegian Barents Sea (Division IIa) bottom trawl survey (February) from 1986-2010 (joint with Russia some of the years since 2000) in fishing depths of $100-500 \mathrm{~m}$. Length compositions for the years 1986-2010 are shown in Table D12a and Fig 7.4a. Age compositions for the years 1992-2008 are shown in Table D12b and Figure 7.4b. This survey covers important nursery areas for the stock.

2 ) Norwegian Svalbard (Division Ilb) bottom trawl survey (AugustSeptember) from 1985-2009 in fishing depths of 100-500 m (depths down to 800 m incl. in the swept area). Length compositions for the years 19852009 and age compositions for the years 1992-2008 are shown in Table D13a and D13b, respectively. This survey covers the northernmost part of
the species' distribution. Insufficient number of age readings in 2009 did not allow for updating the age composition in 2009.
3 ) Data on length and age from both these surveys have been combined and are shown in Figures 7.5a,b.
4 ) Age disaggregated catch rates (numbers $/ \mathrm{nm}^{2}$ averaged for all stations within subareas and finally averaged, weighted by subarea, for the total surveyed area) of Sebastes marinus from the Norwegian Coastal and Fjord survey in 1995-2009 from Finnmark to Møre (Table D14). The series was updated from last year's assessment for 2009. Observations in 2009 indicate maximum catch rates for the $35-44 \mathrm{~cm}$ length group, as before. The estimated catch rates in 2009 were particular high due to one trawl station with an exceptional high catch.
5 ) The bottom trawl surveys covering the Barents Sea and the Svalbard areas show that the abundance indices over the commercial size range ( $>25 \mathrm{~cm}$ ) were relatively stable up to 1998 but declined to lower levels afterwards. Abundance of pre-recruits ( $<25 \mathrm{~cm}$ ) has steadily decreased since 1986 and has remained at very low levels since 2000 (Fig 7.4a).
Results from the Norwegian Coastal and Fjord survey confirm poor recruitment up to 2008. Variation in the results from year to year may be due to a variable number of trawl stations taken in some of the areas from year to year, and annual variations in local fish migrations (Table D14). The distribution of S. marinus is spatially very clustered and the catch rates-at-length estimates are sensitive to few (or even one) station where catches are high. The sharp increase in 2009 should hence be interpreted with great caution (see next chapter).

### 7.3 Assessment with the GADGET model

## Description of the model

Since AFWG2005, experimental analytical assessments have been conducted on this stock using GADGET, and results presented for the years 1990 - last year.

The GADGET model used for the assessment of S. marinus in areas I and II is closely related to the GADGET model that currently is used by the ICES North-Western WG on S. marinus (Björnsson and Sigurdsson 2003). The functioning of a Gadget model, including parameter estimation and data used for tuning, is described in Bogstad et al. (2004) and in the latest Quality Handbook for S. marinus (2009).

## Data used for tuning

Quarterly length distribution of total international commercial landings from two commercial fishing fleets, i.e., Norwegian gillnet and 'all others'. Due to late data submissions, there is one year time lag in the inclusion of length distributions from other countries than Norway.

Quarterly age-length keys from the same fishing fleets, up to 2008
Length disaggregated survey indices from the Barents Sea (Division IIa) bottom trawl survey (February) from 1990-2009 (Table D12a),
Age-length keys and aggregated survey indices from the same survey up to 2009 (Table D12b),

Length disaggregated catch rates (numbers/nautical mile) of Sebastes marinus from the Norwegian Coastal and Fjord survey in 1995-2007 from Finnmark to Møre (Division IIa) (Table D14).

## Changes made to the model and in input data compared with last year's Working Group

Model configuration and settings are identical to that of 2008. Commercial catch data have been revised for years 2008 and updated with year 2009. The proportion mature data set has been extended to 2009, and as noted above the period 1986-1992 has been populated using the average of 1993-1995 in order to avoid model artefacts. The data from 1990-1992 had previously been removed (AFWG 2009) as it was inconsistent with later data. This change has not altered the overall biomass in the model.

A difference was discovered between the Gadget input data for the Barents Sea February survey in 1998 and the data given in Tables D12a,b. The reason for this is that the input data to Gadget had not been corrected for a very large catch on a single station. The input data were corrected to be in accordance with the data given in the tables.

In the previous AFWG the coastal survey length distribution series was only included up to 2005, due to an error in data processing. It can be seen (Figure 7.6) that there is a significant residual pattern in the fit between the model and the survey index, with 2008 and 2009 in particular being very much above the modelled trend. In terms of the length distribution within the survey, the recent years (especially 2008 and 2009) of this survey show the presence of significantly more large redfish than previously (table D14). This trend is not seen in the winter survey (table D12a), and the presence of such large numbers of old fish is not consistent with earlier years of the coastal survey. As a consequence it was decided to exclude 2008 and 2009 of the coastal survey from the model. Furthermore the weighting procedure used on the data sets within the model aims to prevent any single data set from dominating the model fit. As a result of the increasing misfit the coastal survey has been downweighted by approximately $1 / 3$ in order to keep its overall contribution to the fit the same as in previous years.

Experimenting with including the 2008 and 2009 coastal survey data resulted in many more mature fish in the model throughout the time period in order to have these available at to be surveyed in 2008 and 2009, and higher residuals to the other data. The model also produced a higher overall biomass as a result, and consequently slightly low er values for $F$. The alterations in the reasons for the changing signal from the coastal survey in recent years should be investigated inter-sessionally, and should be considered at the forthcoming benchmark workshop for S.marinus.

## Assessment results using the Gadget model

The text table below compares the results from this year's Gadget model with the four previous years.

|  | Totalstock (3+) by 1 January 1990 (tons) | Mean weight in stock 1990 (kg) | SSB (15+) by <br> 1 January <br> $1990^{1}$ (tons) | Totalstock (3+) by 1 January 2003 (tons) | Mean weight in stock 2003 (kg) | SSB (15+) by <br> 1 January <br> $2003^{1}$ (tons) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { WG } \\ & 2006 \end{aligned}$ | 179313 | 0.39 | 64019 | 71013 | 0.71 | 38927 |
| $\begin{aligned} & \text { WG } \\ & 2007 \end{aligned}$ | 163536 | 0.35 | 66712 | 64240 | 0.64 | 43096 |
| $\begin{aligned} & \text { WG } \\ & 2008 \end{aligned}$ | 158851 | 0.35 | 64838 | 74717 | 0.78 | 47693 |
| $\begin{gathered} \text { WG } \\ 2009 \end{gathered}$ | 149763 | 0.34 | 66153 | 73673 | 0.77 | 51683 |
| $\begin{aligned} & \text { WG } \\ & 2010 \end{aligned}$ | 152419 | 0.34 | 58774 | 80073 | 0.79 | 55995 |

${ }^{1)}$ Since WG2007 based on modeled maturation and not 15+, data series used for estimation of maturity modified in 2010

The general patterns in the stock dynamics of S. marinus are similar to those modelled in 2009, with several minor changes. The improvement of the maturity at age data in the early part of the time series resulted in less mature biomass and more immature biomass (but unchanged total biomass) prior to 1995. It is likely that the new pattern is more realistic. The addition of the most recent data, and especially the extension of the coastal survey to 2007, has resulted in a slight increase in the abundance and biomass estimates throughout the time series. This increase is mostly in the mature biomass, and results from signals in the coastal survey. Furthermore the estimated recruitment is revised slightly upwards in prior to 2001 and downwards in more recent years is revised downwards. However, all of these changes are relatively minor, and the overall picture of the stock development remains unchanged from last year.

The most important conclusions to be drawn from the current assessment using the Gadget model are:

- The recruitment to the stock is very poor (Figure 7.9) but increasing, although estimated abundance for new year classes arehighly uncertain.
- The estimated fishing mortality has declined since 1990 and increased again since 2005. The current mortality is estimated around 0.15 (Figure 7.8).
- According to the model the total stock biomass (3+) of S. marinus has decreased from about 150,000 tonnes in 1992-1993 to less than 45,000 tonnes in 2009 (Figure 7.10, Table 7.8).
- The spawning stock biomass of S. marinus has decreased from a maximum of about 70 thousand tonnes in 1996 to approximately 32 thousand tonnes in 2009 ( $-54 \%$, Figure 7.10, Table 7.8). The spawning stock in numbers (SSN) is declining faster than spawning stock biomass (SSB). This is primarily the result of low recruitment in the last $10-15$ years.
- The new treatment of the maturity data has improved the stability of the modelled historical maturity ogives.
- There is increasing uncertainty due to discrepancies in the signals from the different surveys, which needs further investigation


### 7.4 State of the stock

Survey observations and Gadget assessment update confirm previous diagnostics that this stock is currently in a very poor situation. This situation is expected to remain for several years irrespective of current management actions. Year-classes recruit in the SSB at old age ( $\sim 12$ years) and surveys indicate failure of recruitment over a long period. There are indications that new recruits $(<15 \mathrm{~cm})$ may have entered the population in recent years as noted in previous AFWG reports. However it is not clear if this trend genuinely reflects increased S. marinus recruitment, or if it results from species misidentification (with S. mentella).
The analytical assessment using the Gadget model confirms the poor stock situation, and quantifies the development of this stock during the last decade. It is also meant to be an aid for managers to better quantify necessary stronger regulations.
Clearly the stock has at present a reduced reproductive potential and the model suggests that the declining trend in biomass is still going on. In order to reverse this negative development, no directed fishery should be conducted on this stock until a clear increase in the number of juveniles has been detected in surveys, and an improved situation of the mature stock is confirmed by the assessment.

The divergence in the signal in recent years between the two surveys employed here increases the uncertainty in assessing the state of this stock. Further investigation is required to reduce this uncertainty. How ever it should be noted that this uncertainty does not affect the overall conclusions of a continuing decline in biomass, and poor recent recruitment.

Sebastes marinus is currently on the Norwegian Redlist as a vulnerable (VU) species according to the criteria given by the International Union for Conservation of Nature (IUCN). The Royal Norwegian Ministry of Fisheries and Coastal Affairs asked ICES in 2009 to undertake an evaluation of the IUCN criteria used for redlisting marine fish species.
Redlisting is understood to mean a species (or stock) is at risk of extinction. This advice is relevant to Norway's domestic redlisting process. ICES convened two workshops in 2009. The first Workshop (WKPOOR1 (ICES, 2009a)) addressed methods for evaluating extinction risk, and outlined approaches that could support advice on how to avoid potential extinction. The second Workshop (WKPOOR2 (ICES, 2009b)) applied the results of the first workshop to four stocks selected as being of interest to Norway and ICES.
There are three general methods for evaluating extinction risk: (1) screening methods, such as the IUCN redlisting criteria; (2) simple population viability analysis (PVA) based on time trends; and (3) age structured population viability analysis. None of the methods are considered reliable for accurately estimating the absolute probability of extinction, but they may be useful to evaluate the relative probability of extinction between species or between management options.
Simulations were performed on the Sebastes marinus stock using the assumption that the poor recruitment observed during the 1999-2002 period (an average of 26.8 million recruits) would apply in the future, with recruitment independent of the spawning biomass. Simulations done by WKPOOR2 indicate that a constant catch above about 6500 tonnes will lead to a progressive reduction of the stock, and a collapse within 10-15 years if recruitment remains low. However, small changes in recruitment and other parameters that enter the assessment will alter these limits. Neverthe-
less, it seems clear that the current level of catches is at best marginal, and most likely will lead to a stock collapse without a substantial increase in recruitment.

### 7.5 Comments on the Ass essment

The current model assumes constant selectivity through time. It may be possible to extend this to allow for varying selectivity. The model may also be used for comparing modeled mean length at age with the actual data as a contribution to the age reading validation.
S. marinus is considered to be an easier species to age than S. mentella, and it is possible to follow year classes through the input survey data series. An annual updated database on catch-in-numbers at age and length, weight-at-age, and trawl survey indices both by length and age should be continued to be used in future assessment methods.

The current DEEPFISHMAN EU-funded project will aim to use a Gadget S. mentella model as an operating model to assess different simpler assessment methodologies. The approach, if successful, may have implications for producing a simplified assessment model for $S$. marinus.

Further investigation is required into the changing signal from the coastal survey. In addition it is unclear to what extent the slight increase in recruitment in recent years is genuine $S$. marinus recruitment, and how much is due to species misidentification.

### 7.6 Biological reference points

Until an analytical assessment can be accepted and used as basis for reference points calculations for this stock, candidate reference points for the biomass could be set at the average biomass level, or at a certain percentage of this level, estimated by the Russian and Norwegian trawl surveys since 1986. ACFM is supporting this suggestion and states that U-type reference points could be developed provided that a sufficient long time series demonstrating a dynamic range is available. Also the reference point should be expressed in biomass units (SSB or fishable stock), and work has hence been initiated to present the survey time series also in biomass units (also as SSB and fishable stock).

A maximum exploitation rate of $5 \%$ has been suggested sustainable for long lived species like Sebastes spp. when the stocks show no sign of reduced reproductive potential (ref. pelagic redfish in the Irminger Sea and for several rockfishes in the Pacific). Based on the selection curves for the fleets, a reasonable classification of the fishable biomass would be the mature biomass. A corresponding 5\% harvest of this would yield not more than $1,600 \mathrm{t}$, which is well below the current landings and those expected for 2010 of around 6,000 $t$.

### 7.7 Management advice

AFWG considers that the area closures and low bycatch limits should be retained, but stronger regulations than those recently enforced are needed given the continued decline in SSB and low recruitment. Despite the extended ban on the directed fishery by conventional gears from 3 months in 2006 to 5 months in 2007, the current measures are considered insufficient measures to stop the stock from declining to such low levels that any S. marinus fisheries in future will be difficult to conduct. More stringent protective measures should thus be implemented. No directed fishery should be
conducted on this stock at the moment, and the percent legal bycatch should be set as low as possible for other fisheries to continue.

### 7.8 Implementing the ICES Fmsy framework

It should be noted that the current fishery ( $\mathrm{F}=0.15$ ) is well above the suggested $\mathrm{F}_{\mathrm{pa}}$ of $5 \%$ of the stock (Section 7.6). The initial focus should therefore be on reducing total F to no higher than $\mathrm{F}_{\mathrm{pa}}$.

During the ICES W orkshop on Implementing the ICES Fmsy framework (WKFRAME), the closely related beaked redfish Sebastes mentella stock in Sub-areas I and II was used as a case study (ICES 2010) for a data limited situation. The results of this W orkshop refer also to Sebastes marinus in the Barents Sea, where the AFWG is faced with a data limited situation. WKFRAME recommends that the bounds for Fmsy proxies should be evaluated in function of the YPR and SPR curves, and that the reproductive capacity of the $S$. mentella (in this case S. marinus) stock be at least above $30 \%$ of the SPR at $\mathrm{F}=0$. The YPR curve left of the plateau can be used as low bound ( $\mathrm{F}_{01}$ proxy) and a prescribed per-cent SPR as upper bound. The WKFRAME also illustrates by examples why it is informative and important to carry out sensitivity analyses, particularly assumptions regarding natural mortality, selection pattern, growth (density dependence) and maturity.

The AFWG supports the above recommendation by WKFRAME, and that spawner per recruit curves should be provided. The WG did some preliminary estimations ( $\mathrm{F}_{0.1}$ and $\mathrm{F}_{5 P R 40 \%}$ in the order of 0.09-0.12), but recommends that this should be part of the intersessional work until next year's assessment and the proposed benchmark assessment in 2012, including improving the input data for such calculations and evaluating the most appropriate SPR level to be used as reference point for the management of this stock. Evaluations of long lived species with relatively low productivity such as rockfish (Sebastes spp) in the Pacific west coast, concluded that higher SPR values $(50 \%$ to $60 \%)$ were required to maintain sustainable exploitation of these stocks (e.g., Dorn 2002). In the case of S. marinus preliminary estimations resulted in $\mathrm{FSPR} 5 \%=0.08$ and $\mathrm{FSPR}^{2} 0 \%=0.06$.

### 7.9 Response to RGAFNW Technical Minutes

Concerning accuracy and precision of redfish age reading, the AFWG refers to the ICES Redfish age reading workshops in 2005 and 2008 (ICES CM 2006/RMC:09. ICES 2010/xxx) and Stransky et al. (2005) which also recommends how to proceed with age reading of Sebastes spp. for assessment purpose.

The AFWG is convinced that accurate and precise age reading of Sebastes marinus is possible provided that agreed procedures are followed and necessary focus and labour is put into this important basic work for stock assessments. An implementation of QA/QC in the different laboratories involved in age reading of redfish needs to be done, and for stock assessment and regular precision monitoring, a confidence index is proposed. Intercalibration of redfish ageing is urgently needed in order to provide consistent input data for stock assessment. At present, age reading of S. marinus in Sub-areas I and II is only conducted by Norway on a routine basis and for assessment purpose. A high quality assessment of this stock in future is completely dependent on that the age reading is continued. Proper quality assurance of the age reading is dependent on having more than one reader, and regular intercalibrations among national and international readers should be conducted.

There is a lack of data to directly estimate the natural mortality (M) for S. marinus. The WG has applied a $\mathrm{M}=0.1$, which has been considered suitable for a long lived species such as $S$. marinus. For Sebastes marinus in ICES sub-areas V and XIV the NWWG has set the natural mortality to 0.15 for the youngest age, decreasing gradually to 0.05 for age 5 and older. In the Pacific, e.g., Dorn (2002) presents an overview of the different natural mortalities used for different Sebastes species in that area. The WG decided not to change the long term practice of setting this to 0.1 until this has been better investigated, e.g., in connection with a benchmark assessment.

With the exception of a slight increase in the catchability of small individuals (less than $18-20 \mathrm{~cm}$ ) in the trawl survey since 1993, there is no evidence of changes of catchability. How ever the WG will continue to monitor the catchability data, and retains the technical ability to implement varying catchability in the model if the evidence suggests it is needed.
The review group shares the view of the AFWG and stated that a benchmark assessment is needed for this stock (expected in 2012). Until then, due to the expected low recruitment, the review group further recommended that the advice for this stock can be based on the current assessment method.

Table 7.1 Sebastes marinus in Sub-areas I and II. Nominal catch (t) by countries in Sub-area I and Divisions IIa and IIb combined.

| Year | Fa roe Islands | France | Germany ${ }^{2}$ | Greenland | Ice land | Ire land | Ne the rlands |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 3 | 796 | 412 | - | - | - | - |
| 1990 | 278 | 1,679 | 387 | 1 | - | - | - |
| 1991 | 152 | 706 | 981 | - | - | - | - |
| 1992 | 35 | 1,289 | 530 | 623 | - | - | - |
| 1993 | 139 | 871 | 650 | 14 | - | - | - |
| 1994 | 22 | 697 | 1,008 | 5 | 4 | - | - |
| 1995 | 27 | 732 | 517 | 5 | 1 | 1 | 1 |
| 1996 | 38 | 671 | 499 | 34 | - | - | - |
| 1997 | 3 | 974 | 457 | 23 | - | 5 | - |
| 1998 | 78 | 494 | 131 | 33 | - | 19 | - |
| 1999 | 35 | 35 | 228 | 47 | 14 | 7 | - |
| 2000 | 17 | 13 | 160 | 22 | 16 | - | - |
| 2001 | 37 | 30 | 238 | 17 | - | 1 | - |
| 2002 | 60 | 31 | 42 | 31 | 3 | - | - |
| 2003 | 109 | 8 | 122 | 36 | 4 | - | 89 |
| 2004 | 19 | 4 | 68 | 20 | 30 | - | 33 |
| 2005 | 47 | 10 | 72 | 36 | 8 | - | 48 |
| 2006 | 111 | 8 | 35 | 44 | 31 | 3 | 21 |
| 2007 | 146 | 15 | 67 | 55 | 68 | 1 | 20 |
| 2008 | 274 | 63 | 30 | 49 | 27 | - | 2 |
| $2009{ }^{1}$ | 70 | 4 | 58 | 27 | 13 | - | 1 |
| Year | Norway | Portugal | Russia ${ }^{3}$ | Spain | UK (Eng. \& | UK | Total |
|  |  |  |  |  | Wales) | (Scotl) |  |
| 1989 | 20,662 | - | 1,264 | - | 97 | ( | 23,234 |
| 1990 | 23,917 | - | 1,549 | - | 261 | - | 28,072 |
| 1991 | 15,872 | - | 1.052 | - | 268 | 10 | 19,041 |
| 1992 | 12,700 | 5 | 758 | 2 | 241 | 2 | 16,185 |
| 1993 | 13,137 | 77 | 1,313 | 8 | 441 | 1 | 16,651 |
| 1994 | 14,955 | 90 | 1,199 | 4 | 135 | 1 | 18,120 |
| 1995 | 13,516 | 9 | 639 | - | 159 | 9 | 15,616 |
| 1996 | 15,622 | 55 | 716 | 81 | 229 | 98 | 18,043 |
| 1997 | 14,182 | 61 | 1,584 | 36 | 164 | 22 | 17,511 |
| 1998 | 16,540 | 6 | 1,632 | 51 | 118 | 53 | 19,155 |
| 1999 | 16,750 | 3 | 1,691 | 7 | 135 | 34 | 18,986 |
| 2000 | 13,032 | 16 | 1,112 | - |  | $73^{4}$ | 14,461 |
| 2001 | 9,134 | 7 | 963 | 1 |  | 1194 | 10,547 |
| 2002 | 8,561 | 34 | 832 | 3 |  | $46^{4}$ | 9,643 |
| 2003 | 6,853 | 6 | 479 | - |  | $134{ }^{4}$ | 7,840 |
| 2004 | 6,233 | 5 | 722 | 3 |  | $69^{4}$ | 7,206 |
| 2005 | 6,085 ${ }^{1}$ | 56 | 614 | 8 |  | $52^{4}$ | 7,037 |
| 2006 | 6,265 ${ }^{1}$ | 69 | 713 | 9 |  | $39^{4}$ | 7,348 |
| 2007 | 5,7591 | 225 | 890 | 5 |  | $55^{4}$ | 7,306 |
| 2008 | 5,202 | 72 | 749 | 4 |  | 85 | 6,557 |
| $2009{ }^{1}$ | 5,383 | 30 | 698 | - |  | 9 | 6,293 |

[^7]Table 7.2 Sebastes marinus. Nominal catch (t) by countries in Sub-area I.

| Year | Fa roe Islands | Germany ${ }^{4}$ | Greenland | Ice land | Norway | Russia ${ }^{5}$ | UK(Eng\&W ales) | UK(Scotl) | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | - | - | - | - | 1,763 | 110 | $4^{2}$ | - | 1,877 |
| 1990 | 5 | - | - | - | 1,263 | 14 | - | - | 1,282 |
| 1991 | - | - | - | - | 1,993 | 92 | - | - | 2,085 |
| 1992 | - | - | - | - | 2,162 | 174 | - | - | 2,336 |
| 1993 | $24^{2}$ | - | - | - | 1,178 | 330 | - | - | 1,532 |
| 1994 | $12^{2}$ | 72 | - | 4 | 1,607 | 109 |  | - | 1,804 |
| 1995 | $19^{2}$ | $1{ }^{2}$ | - | $1{ }^{2}$ | 1,947 | 201 | $1^{2}$ | - | 2,170 |
| 1996 | $7^{2}$ | - | - | - | 2,245 | 131 | $3^{2}$ | - | 2,386 |
| 1997 | $3^{2}$ | - | $5^{2}$ | - | 2,431 | 160 | $2^{2}$ | - | 2,601 |
| 1998 | $78^{2}$ | $5^{2}$ | - | - | 2,109 | 308 | $30^{2}$ | - | 2,530 |
| 1999 | $35^{2}$ | $18^{2}$ | $9^{2}$ | $14^{2}$ | 2,114 | 360 | $11^{2}$ | - | 2,561 |
| 2000 | - | $1^{2}$ | - | $16^{2}$ | 1,983 | 146 |  | $12^{6}$ | 2,159 |
| 2001 | 4 | $11^{2}$ | - | - | 1,053 | 128 | France | $16^{6}$ | 1,212 |
| 2002 | 15 | $5^{2}$ | - | - | 693 | 220 | $1{ }^{2}$ | 92,6 | 943 |
| 2003 | $15^{2}$ | - | 1 | - | 815 | 140 | - | $4^{6}$ | 975 |
| 2004 | 7 | - | - | - | 1,237 | 213 | - | $12^{6}$ | 1,469 |
| 2005 | 10 | - | - | - | 1,002 ${ }^{1}$ | 61 | 1 | $4^{6}$ | 1,078 |
| 2006 | 46 | - | - | - | 685 | 136 | - | - | 867 |
| 2007 | 15 | 12 | Spain- 2 | - | 1,029 | 49 | - | $20^{6}$ | 1,127 |
| 2008 | 45 | 2 | Portug- 3 | Ltu- | 632 | 49 | 7 | 15 | 754 |
| $2009{ }^{1}$ | - | $3{ }^{2}$ | 13 | - | 678 | 19 | - | - | 713 |

## 1 Provisional figures.

2 Split on species according to reports to Norwegian authorities.
3 Based on preliminary estimates of species breakdown by area.
4 Includes former GDR prior to 1991.
5 USSR prior to 1991.

## 6 UK(E\&W)+UK(Scot.)

7 Split on species according to reports to Russian authorities.

Table 7.3 Sebastes marinus. Nominal catch (t) by countries in Division IIa.

| Year | Fa roe F Islands | France | Germany ${ }^{4}$ | Greenland | Ire- <br> land | Ne ther- <br> lands | Norway | Portugal | Russia ${ }^{5}$ S | pain | (Eng. \& Wales) | $\begin{gathered} \text { UK } \\ (\text { Scotl.) } \end{gathered}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | $3{ }^{2}$ | $784{ }^{2}$ | 412 | - | - | - | 18,833 | - | 912 |  | $93^{2}$ | - | 21,037 |
| 1990 | 273 | 1,684 ${ }^{2}$ | 387 | - |  | - - | 22,444 | - | 392 |  | 261 | - | 25,441 |
| 1991 | $152{ }^{2}$ | $706^{2}$ | 678 | - |  | - | 13,835 | - | 534 |  | $268{ }^{2}$ | $10^{2}$ | 16,183 |
| 1992 | $35^{2}$ | 1,294 ${ }^{2}$ | 211 | 614 | - | - - | 10,536 | - | 404 |  | $206{ }^{2}$ | $2^{2}$ | 13,302 |
| 1993 | $115^{2}$ | $871^{2}$ | 473 | $14^{2}$ | - | - - | 11,959 | $77^{2}$ | 940 |  | $431^{2}$ | $1{ }^{2}$ | 14,881 |
| 1994 | $10^{2}$ | $697{ }^{2}$ | $654{ }^{2}$ | $5^{2}$ | - | - | 13,330 | $90^{2}$ | 1,030 |  | 1292 | - | 15,945 |
| 1995 | $8^{2}$ | $732{ }^{2}$ | $328{ }^{2}$ | $5^{2}$ | $1{ }^{2}$ | 1 | 11,466 | $2^{2}$ | 405 |  | $158{ }^{2}$ | $9{ }^{2}$ | 13,115 |
| 1996 | $27^{2}$ | $671{ }^{2}$ | $448{ }^{2}$ | $34^{2}$ | - | - | 13,329 | $51^{2}$ | 449 | $5^{2}$ | $22^{2} 3^{2}$ | $98^{2}$ | 15,335 |
| 1997 | - | $974{ }^{2}$ | 438 | $18^{2}$ | $5^{2}$ | - | 11,708 | $61^{2}$ | 1,199 | $36^{2}$ | 2 $162^{2}$ | $22^{2}$ | 14,623 |
| 1998 | - | $494{ }^{2}$ | $116^{2}$ | $33^{2}$ | $19^{2}$ | - | 14,326 | $6^{2}$ | 1,078 | $51^{2}$ | $25^{2}$ | $52^{2}$ | 16,260 |
| 1999 | - | $35^{2}$ | $210^{2}$ | $38^{2}$ | $7^{2}$ | - | 14,598 | $3^{2}$ | 976 | $7^{2}$ | $122^{2}$ | $34^{2}$ | 16,030 |
| 2000 | $17^{2}$ | $13^{2}$ | $159{ }^{2}$ | $22^{2}$ | - | - | 11,038 | $16^{2}$ | 658 |  | - | 61 | 11,984 |
| 2001 | $33^{2}$ | $30^{2}$ | $227^{2}$ | $17^{2}$ | $1^{2}$ | - | 8,002 | $6^{2}$ | 612 |  | 2 Iceland | $103{ }^{2}$ | 9,031 |
| 2002 | $45^{2}$ | $30^{2}$ | $37^{2}$ | $31^{2}$ | - | - - | 7,761 | $18^{2}$ | 192 | $2^{2}$ | $23^{2}$ | $32^{2}$ | 8,151 |
| 2003 | $94^{2}$ | $9{ }^{2}$ | $122^{2}$ | $35^{2}$ | - | $89^{2}$ | 5,970 | $6^{2}$ | 264 |  | $4^{2}$ | $130^{2}$ | 6,722 |
| 2004 | $12^{2}$ | $4^{2}$ | $68^{2}$ | $20^{2}$ | - | $33^{2}$ | 4,872 | $5^{2}$ | 396 | $3^{2}$ | $20^{2}$ | $58^{2}$ | 5,500 |
| 2005 | $37^{2}$ | $9^{2}$ | $60^{2}$ | $36^{2}$ | - | 48 | 4,855 ${ }^{1}$ | $56^{2}$ | 265 | $8^{2}$ | $8^{2} \quad 8^{2}$ | $48^{2}$ | 5,430 |
| 2006 | $60^{2}$ | $8^{2}$ | $35^{2}$ | $44^{2}$ | $3^{2}$ | $21^{2}$ | 4,404 | $59^{2}$ | 293 | $9{ }^{2}$ | $231{ }^{2}$ | $39^{2}$ | 5,006 |
| 2007 | $119{ }^{2}$ | $15^{2}$ | $55^{2}$ | $55^{2}$ | $1{ }^{2}$ | $20^{2}$ | 4,101 ${ }^{1}$ | 70 | 599 | $3^{2}$ | 268 | $35^{2}$ | 5,142 |
| 2008 | $229{ }^{2}$ | $56^{2}$ | $28^{2}$ | $49^{2}$ | - | $2^{2}$ | 4,444 | $68^{2}$ | 450 | $4^{2}$ | $27^{2}$ | $70^{2}$ | 5,426 |
| $2009{ }^{1}$ | $70^{2}$ | $4^{2}$ | $55^{2}$ | $27^{2}$ | - | $1^{2}$ | 4,504 | $17^{2}$ | 500 |  | $13^{2}$ | $9^{2}$ | 5,200 |

${ }^{1}$ Provisional figures.
${ }^{2}$ Split on spe cies a ccording to reports to Norwe gian authorities.
${ }^{3}$ Base d on pre limina ry estimates of species breakdown by a rea.
${ }^{4}$ Includes former GDR prior to 1991.
${ }^{5}$ USSR prior to 1991.
${ }^{6}$ UK(E\&W) + UK (Scot.)

Table 7.4 Sebastes marinus. Nominal catch (t) by countries in Division IIb.

| Year | Fa roe Islands | Germany ${ }^{5}$ | Greenland | Norway | Portugal | Russia ${ }^{6}$ | Spain | UK(Eng. \& Wales) |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | - | - | - | 66 | - | 242 | - | - | - | 308 |
| 1990 | - | - | $1{ }^{2}$ | 210 | - | 1157 | - | - | - | 1,368 |
| 1991 | - | 303 | - | 44 | - | 426 | - | - | - | 773 |
| 1992 | - | 319 | $9^{2}$ | 2 | $5^{2}$ | 180 | 2 | $35^{2}$ | - | 552 |
| 1993 | - | 177 | - | - | - | 43 | $8^{3}$ | $10^{2}$ | - | 238 |
| 1994 | - | 282 | - | 18 | - | 60 | $4^{3}$ | $6^{2}$ | $1{ }^{2}$ | 371 |
| 1995 | - | 187 | - | 103 | 7 | 33 | - | - | - | 330 |
| 1996 | 4 | $51^{2}$ | - | 27 | 5 | 136 | $76^{2}$ | $3^{2}$ | - | 302 |
| 1997 | - | 20 | - | 43 | - | 225 | - | - | - | 288 |
| 1998 | - | $10^{2}$ | - | 105 | - | 246 | - | $3^{2}$ | - | 364 |
| 1999 | - | - | - | 38 | - | 355 | - | $2^{2}$ | - | 395 |
| 2000 | - | - | - | 10 | - | 308 | - | - | - | 318 |
| 2001 | - | - | - | 79 | $1{ }^{2}$ | 223 | - | - | - | 303 |
| 2002 | - | - | - | 107 | $16^{2}$ | 420 | $1{ }^{2}$ | - | $5^{2,7}$ | 549 |
| 2003 | - | - | - | 68 | - | 75 | - | - | - | 143 |
| 2004 | - | - | - | 124 | - | 113 | - | - | - | 237 |
| 2005 | - | $13^{2}$ | - | $228{ }^{1}$ | - | 288 | - | - | - | 529 |
| 2006 | $5^{2}$ | - | - | 1,211 ${ }^{1}$ | $10^{2}$ | 284 | - | - | - | 1,510 |
| 2007 | 122 | - | - | 649 | 155 | 242 | - | - | - | 1,057 |
| 2008 | - | - | - | 126 | $1{ }^{2}$ | 250 | - | - | - | 377 |
| $2009^{1}$ | - | - | - | 200 | - | 179 | - | - | - | 379 |

1 Provisional figures.
2 Split on species according to re ports to Norwe gianauthorities.
3 Split on species according to the 1992 catches.
4 Based on prelimina ry estimates of species breakdown by a rea.
5 Includes former GDR prior to 1991.
6 USSR prior to 1991.
7UK(E\&W)+UK(Scot.)

Table 7.5. Sebastes marinus in Sub-areas I and II. Catch numbers at age (in thousands).

| Year/Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7 | 0 | 46 | 60 | 9 | 9 | 28 | 78 | 4 | 23 | 14 | 22 | 19 | 40 | 45 | 15 | 1 | 0 |
| 8 | 24 | 7 | 85 | 119 | 98 | 51 | 593 | 13 | 23 | 36 | 25 | 47 | 55 | 32 | 21 | 4 | 0 |
| 9 | 193 | 292 | 230 | 313 | 156 | 206 | 855 | 70 | 44 | 71 | 30 | 46 | 94 | 56 | 31 | 14 | 6 |
| 10 | 359 | 640 | 672 | 361 | 321 | 470 | 572 | 245 | 199 | 143 | 44 | 65 | 80 | 70 | 68 | 12 | 5 |
| 11 | 406 | 816 | 908 | 879 | 686 | 721 | 1006 | 902 | 347 | 414 | 204 | 198 | 165 | 245 | 138 | 49 | 16 |
| 12 | 1036 | 1930 | 1610 | 1234 | 1065 | 968 | 1230 | 958 | 482 | 686 | 359 | 277 | 173 | 204 | 306 | 139 | 18 |
| 13 | 1022 | 2096 | 2038 | 1638 | 1781 | 1512 | 1618 | 1782 | 1120 | 1199 | 705 | 504 | 393 | 201 | 448 | 265 | 110 |
| 14 | 1523 | 2030 | 2295 | 2134 | 2276 | 1736 | 1480 | 1409 | 1342 | 1943 | 1687 | 590 | 779 | 809 | 495 | 366 | 425 |
| 15 | 2353 | 1601 | 1783 | 1675 | 2172 | 1582 | 1612 | 2121 | 1674 | 1377 | 1338 | 677 | 741 | 549 | 523 | 361 | 266 |
| 16 | 1410 | 2725 | 1406 | 1614 | 1848 | 1045 | 1239 | 2203 | 1653 | 1274 | 1071 | 963 | 916 | 779 | 637 | 443 | 230 |
| 17 | 1655 | 2668 | 785 | 1390 | 1421 | 1277 | 1407 | 1715 | 1243 | 1196 | 937 | 1059 | 926 | 794 | 892 | 442 | 454 |
| 18 | 1678 | 1409 | 563 | 952 | 851 | 970 | 1558 | 753 | 568 | 388 | 481 | 787 | 743 | 747 | 616 | 538 | 310 |
| 19 | 745 | 617 | 670 | 679 | 804 | 1018 | 1019 | 483 | 119 | 313 | 367 | 436 | 376 | 496 | 510 | 547 | 591 |
| 20 | 716 | 733 | 593 | 439 | 608 | 846 | 394 | 458 | 183 | 99 | 146 | 169 | 210 | 332 | 396 | 479 | 549 |
| 21 | 534 | 514 | 419 | 560 | 511 | 443 | 197 | 132 | 154 | 104 | 84 | 183 | 189 | 310 | 225 | 281 | 386 |
| 22 | 528 | 256 | 368 | 334 | 205 | 764 | 459 | 230 | 112 | 117 | 51 | 108 | 129 | 188 | 322 | 223 | 229 |
| 23 | 576 | 177 | 250 | 490 | 334 | 486 | 174 | 224 | 135 | 113 | 18 | 79 | 111 | 165 | 170 | 144 | 236 |
| +gp | 3482 | 1508 | 3232 | 3135 | 2131 | 3389 | 2131 | 895 | 254 | 253 | 69 | 186 | 220 | 397 | 630 | 1032 | 696 |
| TOTALNUM | 18240 | 20065 | 17967 | 17955 | 17277 | 17512 | 17622 | 14597 | 9675 | 9740 | 7637 | 6390 | 6338 | 6419 | 6443 | 5342 | 4526 |
| TONSLAND | 16651 | 18120 | 15616 | 18043 | 17511 | 19155 | 18986 | 14460 | 10547 | 9643 | 7841 | 7320 | 7037 | 7,348 | 7306 | 6557 | 6292 |

Table 7.6. Sebastes marinus in Sub-areas I and II. Catch weights at age (kg)

| Year/Age | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{7}$ | 0.20 | 0.25 | 0.33 | 0.22 | 0.23 | 0.37 | 0.14 | 0.19 | 0.15 | 0.17 |
| $\mathbf{8}$ | 0.33 | 0.37 | 0.43 | 0.49 | 0.51 | 0.21 | 0.26 | 0.24 | 0.26 | 0.25 |
| $\mathbf{9}$ | 0.36 | 0.38 | 0.64 | 0.56 | 0.53 | 0.47 | 0.44 | 0.32 | 0.45 | 0.33 |
| $\mathbf{1 0}$ | 0.43 | 0.49 | 0.61 | 0.65 | 0.74 | 0.62 | 0.57 | 0.44 | 0.55 | 0.42 |
| $\mathbf{1 1}$ | 0.51 | 0.51 | 0.59 | 0.71 | 0.72 | 0.67 | 0.69 | 0.53 | 0.58 | 0.54 |
| $\mathbf{1 2}$ | 0.51 | 0.64 | 0.65 | 0.81 | 0.78 | 0.77 | 0.78 | 0.64 | 0.67 | 0.67 |
| $\mathbf{1 3}$ | 0.64 | 0.74 | 0.74 | 0.84 | 0.80 | 0.77 | 0.86 | 0.73 | 0.80 | 0.72 |
| $\mathbf{1 4}$ | 0.64 | 0.76 | 0.79 | 0.88 | 0.86 | 0.85 | 1.04 | 0.84 | 0.89 | 0.84 |
| $\mathbf{1 5}$ | 0.76 | 0.86 | 0.84 | 0.96 | 0.91 | 1.05 | 1.07 | 0.96 | 1.01 | 0.98 |
| $\mathbf{1 6}$ | 0.86 | 0.95 | 0.92 | 1.00 | 0.99 | 0.96 | 1.12 | 1.11 | 1.14 | 1.09 |
| $\mathbf{1 7}$ | 0.89 | 1.03 | 1.12 | 1.02 | 1.16 | 1.25 | 1.18 | 1.25 | 1.33 | 1.20 |
| $\mathbf{1 8}$ | 0.98 | 1.07 | 1.01 | 1.01 | 1.18 | 1.28 | 1.71 | 1.32 | 1.43 | 1.30 |
| $\mathbf{1 9}$ | 1.00 | 1.11 | 1.01 | 1.00 | 1.21 | 1.30 | 1.09 | 1.53 | 1.62 | 1.44 |
| $\mathbf{2 0}$ | 1.03 | 1.16 | 1.21 | 1.03 | 1.34 | 1.23 | 1.18 | 1.06 | 1.60 | 1.78 |
| $\mathbf{2 1}$ | 1.21 | 1.15 | 1.14 | 1.04 | 1.28 | 1.87 | 1.04 | 1.29 | 1.47 | 1.68 |
| $\mathbf{2 2}$ | 1.03 | 1.13 | 1.09 | 1.14 | 1.54 | 1.46 | 1.34 | 1.32 | 2.00 | 1.88 |
| $\mathbf{2 3}$ | 1.20 | 1.02 | 1.30 | 1.09 | 1.19 | 1.73 | 1.18 | 1.12 | 2.70 | 2.12 |
| $\mathbf{+ g p}$ | 1.14 | 1.36 | 1.01 | 1.16 | 1.29 | 1.29 | 1.34 | 1.20 | 2.31 | 1.84 |


| Year/Age | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ | $\mathbf{2 0 0 9}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{7}$ | 0.19 | 0.21 | 0.16 | 0.13 | 0.15 | 0.41 | - |
| $\mathbf{8}$ | 0.22 | 0.26 | 0.21 | 0.15 | 0.21 | 0.55 | 0.28 |
| $\mathbf{9}$ | 0.31 | 0.36 | 0.36 | 0.28 | 0.33 | 0.55 | 0.33 |
| $\mathbf{1 0}$ | 0.39 | 0.45 | 0.45 | 0.41 | 0.39 | 0.57 | 0.57 |
| $\mathbf{1 1}$ | 0.49 | 0.51 | 0.52 | 0.51 | 0.50 | 0.52 | 0.57 |
| $\mathbf{1 2}$ | 0.58 | 0.59 | 0.58 | 0.58 | 0.59 | 0.58 | 0.59 |
| $\mathbf{1 3}$ | 0.69 | 0.68 | 0.68 | 0.66 | 0.65 | 0.65 | 0.76 |
| $\mathbf{1 4}$ | 0.84 | 0.80 | 0.82 | 0.74 | 0.77 | 0.81 | 0.86 |
| $\mathbf{1 5}$ | 0.96 | 0.96 | 0.94 | 0.83 | 0.90 | 0.90 | 0.95 |
| $\mathbf{1 6}$ | 1.05 | 1.07 | 1.03 | 1.00 | 1.00 | 1.07 | 1.07 |
| $\mathbf{1 7}$ | 1.29 | 1.22 | 1.16 | 1.14 | 1.09 | 1.14 | 1.20 |
| $\mathbf{1 8}$ | 1.36 | 1.34 | 1.36 | 1.27 | 1.27 | 1.36 | 1.34 |
| $\mathbf{1 9}$ | 1.65 | 1.57 | 1.46 | 1.39 | 1.42 | 1.51 | 1.43 |
| $\mathbf{2 0}$ | 1.74 | 1.67 | 1.51 | 1.46 | 1.32 | 1.81 | 1.61 |
| $\mathbf{2 1}$ | 2.09 | 1.75 | 1.67 | 1.37 | 1.53 | 1.99 | 1.67 |
| $\mathbf{2 2}$ | 1.85 | 2.09 | 1.91 | 1.47 | 1.47 | 2.01 | 1.93 |
| $\mathbf{2 3}$ | 2.30 | 1.90 | 2.23 | 1.64 | 1.69 | 2.26 | 1.91 |
| $\mathbf{+ g p}$ | 2.38 | 2.04 | 2.27 | 2.03 | 1.81 | 1.93 | 1.60 |

Table 7.7. Sebastes marinus in Sub-areas I and II. Fishing mortalities as estimated by Gadget.

|  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 5 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 6 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.000 | 0.001 | 0.001 | 0.001 | 0.001 | 0.000 |
| 7 | 0.004 | 0.003 | 0.002 | 0.002 | 0.002 | 0.002 | 0.002 | 0.002 | 0.003 | 0.003 | 0.002 |
| 8 | 0.036 | 0.010 | 0.008 | 0.008 | 0.008 | 0.007 | 0.008 | 0.007 | 0.008 | 0.009 | 0.007 |
| 9 | 0.067 | 0.049 | 0.021 | 0.020 | 0.021 | 0.017 | 0.020 | 0.019 | 0.022 | 0.022 | 0.017 |
| 10 | 0.091 | 0.076 | 0.067 | 0.041 | 0.042 | 0.035 | 0.040 | 0.039 | 0.044 | 0.045 | 0.036 |
| 11 | 0.120 | 0.096 | 0.091 | 0.096 | 0.071 | 0.059 | 0.068 | 0.066 | 0.074 | 0.077 | 0.061 |
| 12 | 0.154 | 0.119 | 0.109 | 0.120 | 0.135 | 0.088 | 0.100 | 0.098 | 0.110 | 0.114 | 0.092 |
| 13 | 0.191 | 0.144 | 0.128 | 0.137 | 0.157 | 0.139 | 0.134 | 0.131 | 0.147 | 0.154 | 0.123 |
| 14 | 0.232 | 0.170 | 0.147 | 0.154 | 0.173 | 0.155 | 0.186 | 0.161 | 0.182 | 0.190 | 0.153 |
| 15 | 0.276 | 0.197 | 0.167 | 0.171 | 0.190 | 0.167 | 0.200 | 0.204 | 0.212 | 0.223 | 0.179 |
| 16 | 0.320 | 0.225 | 0.186 | 0.187 | 0.205 | 0.179 | 0.211 | 0.215 | 0.250 | 0.250 | 0.201 |
| 17 | 0.365 | 0.252 | 0.205 | 0.203 | 0.220 | 0.189 | 0.221 | 0.224 | 0.260 | 0.282 | 0.218 |
| 18 | 0.387 | 0.279 | 0.223 | 0.218 | 0.233 | 0.199 | 0.231 | 0.231 | 0.267 | 0.289 | 0.238 |
| 19 | 0.408 | 0.292 | 0.240 | 0.232 | 0.245 | 0.208 | 0.239 | 0.238 | 0.274 | 0.295 | 0.242 |
| 20 | 0.428 | 0.304 | 0.248 | 0.244 | 0.257 | 0.216 | 0.246 | 0.244 | 0.279 | 0.301 | 0.246 |
| 21 | 0.447 | 0.315 | 0.256 | 0.250 | 0.266 | 0.222 | 0.253 | 0.250 | 0.284 | 0.305 | 0.249 |
| 22 | 0.464 | 0.326 | 0.263 | 0.255 | 0.271 | 0.228 | 0.258 | 0.254 | 0.288 | 0.309 | 0.251 |
| 23 | 0.478 | 0.335 | 0.269 | 0.260 | 0.275 | 0.231 | 0.263 | 0.258 | 0.292 | 0.312 | 0.253 |
| 24 | 0.491 | 0.343 | 0.274 | 0.264 | 0.278 | 0.233 | 0.265 | 0.261 | 0.295 | 0.314 | 0.255 |
| 25 | 0.502 | 0.350 | 0.279 | 0.268 | 0.282 | 0.235 | 0.266 | 0.262 | 0.297 | 0.317 | 0.256 |
| 26 | 0.510 | 0.356 | 0.283 | 0.271 | 0.284 | 0.237 | 0.268 | 0.263 | 0.298 | 0.318 | 0.257 |
| 27 | 0.517 | 0.360 | 0.286 | 0.274 | 0.287 | 0.239 | 0.269 | 0.264 | 0.299 | 0.319 | 0.258 |
| 28 | 0.522 | 0.364 | 0.289 | 0.276 | 0.289 | 0.240 | 0.271 | 0.265 | 0.300 | 0.320 | 0.259 |
| 29 | 0.526 | 0.366 | 0.291 | 0.278 | 0.290 | 0.241 | 0.272 | 0.266 | 0.300 | 0.320 | 0.259 |
| 30 | 0.530 | 0.370 | 0.294 | 0.281 | 0.293 | 0.243 | 0.274 | 0.267 | 0.301 | 0.321 | 0.260 |
|  | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |  |  |
| 4 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |  |  |
| 5 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |  |  |
| 6 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |  |  |
| 7 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.002 | 0.002 | 0.002 | 0.002 |  |  |
| 8 | 0.005 | 0.005 | 0.004 | 0.004 | 0.004 | 0.005 | 0.006 | 0.005 | 0.006 |  |  |
| 9 | 0.013 | 0.012 | 0.010 | 0.010 | 0.011 | 0.013 | 0.014 | 0.014 | 0.015 |  |  |
| 10 | 0.026 | 0.025 | 0.021 | 0.021 | 0.022 | 0.026 | 0.028 | 0.028 | 0.030 |  |  |
| 11 | 0.045 | 0.043 | 0.036 | 0.035 | 0.037 | 0.044 | 0.048 | 0.048 | 0.052 |  |  |
| 12 | 0.068 | 0.064 | 0.054 | 0.053 | 0.055 | 0.064 | 0.071 | 0.071 | 0.079 |  |  |
| 13 | 0.092 | 0.086 | 0.072 | 0.071 | 0.073 | 0.085 | 0.094 | 0.095 | 0.106 |  |  |
| 14 | 0.114 | 0.106 | 0.089 | 0.087 | 0.090 | 0.103 | 0.115 | 0.117 | 0.132 |  |  |
| 15 | 0.134 | 0.123 | 0.104 | 0.101 | 0.103 | 0.118 | 0.132 | 0.136 | 0.154 |  |  |
| 16 | 0.151 | 0.137 | 0.116 | 0.112 | 0.114 | 0.130 | 0.146 | 0.151 | 0.171 |  |  |
| 17 | 0.164 | 0.148 | 0.125 | 0.120 | 0.122 | 0.139 | 0.156 | 0.162 | 0.184 |  |  |
| 18 | 0.173 | 0.157 | 0.132 | 0.126 | 0.128 | 0.145 | 0.163 | 0.170 | 0.194 |  |  |
| 19 | 0.184 | 0.163 | 0.137 | 0.131 | 0.132 | 0.150 | 0.168 | 0.175 | 0.200 |  |  |
| 20 | 0.186 | 0.169 | 0.141 | 0.134 | 0.135 | 0.153 | 0.172 | 0.179 | 0.204 |  |  |
| 21 | 0.188 | 0.170 | 0.144 | 0.136 | 0.137 | 0.155 | 0.174 | 0.181 | 0.207 |  |  |
| 22 | 0.190 | 0.171 | 0.145 | 0.138 | 0.138 | 0.156 | 0.175 | 0.183 | 0.209 |  |  |
| 23 | 0.191 | 0.172 | 0.146 | 0.139 | 0.140 | 0.157 | 0.176 | 0.184 | 0.210 |  |  |
| 24 | 0.192 | 0.173 | 0.146 | 0.139 | 0.140 | 0.158 | 0.177 | 0.185 | 0.211 |  |  |
| 25 | 0.193 | 0.174 | 0.147 | 0.139 | 0.140 | 0.159 | 0.178 | 0.185 | 0.212 |  |  |
| 26 | 0.194 | 0.174 | 0.147 | 0.140 | 0.141 | 0.159 | 0.178 | 0.186 | 0.212 |  |  |
| 27 | 0.194 | 0.175 | 0.147 | 0.140 | 0.141 | 0.159 | 0.178 | 0.186 | 0.212 |  |  |
| 28 | 0.195 | 0.175 | 0.147 | 0.140 | 0.141 | 0.159 | 0.178 | 0.186 | 0.213 |  |  |
| 29 | 0.195 | 0.175 | 0.148 | 0.140 | 0.141 | 0.159 | 0.178 | 0.186 | 0.213 |  |  |
| 30 | 0.195 | 0.176 | 0.148 | 0.141 | 0.141 | 0.159 | 0.179 | 0.186 | 0.213 |  |  |

Table 7.8. Sebastes marinus in Sub-areas I and II. Stock numbers, biomass, mean weight and maturity ogives as estimated by GADGET using two surve y series as input.

| year | Number (millions) | redfish meanweight $(\mathrm{kg})$ | $\begin{aligned} & \text { Biomass } \\ & (1000 \text { 't }) \\ & \hline \end{aligned}$ | Number (millions) | mature meanweight $(\mathrm{kg})$ | $\begin{aligned} & \text { Biomass } \\ & (1000 \text { 't }) \\ & \hline \end{aligned}$ | Number (millions) | $\begin{gathered} \hline \text { immature } \\ \text { meanweight } \\ (\mathrm{kg}) \end{gathered}$ | $\begin{aligned} & \text { Biomass } \\ & (1000 \text { 't }) \\ & \hline \end{aligned}$ | recruit Number (1000') |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 529 | 0.33 | 175 | 114 | 0.79 | 90 | 415 | 0.20 | 85 | 87,615 |
| 1987 | 521 | 0.32 | 167 | 106 | 0.79 | 84 | 414 | 0.20 | 83 | 67,297 |
| 1988 | 499 | 0.33 | 163 | 98 | 0.77 | 76 | 401 | 0.22 | 87 | 51,250 |
| 1989 | 475 | 0.33 | 158 | 90 | 0.74 | 67 | 384 | 0.24 | 91 | 46,525 |
| 1990 | 455 | 0.33 | 152 | 84 | 0.71 | 60 | 371 | 0.25 | 93 | 50,955 |
| 1991 | 443 | 0.34 | 152 | 84 | 0.70 | 59 | 359 | 0.26 | 93 | 50,237 |
| 1992 | 424 | 0.36 | 153 | 86 | 0.71 | 61 | 338 | 0.27 | 92 | 39,665 |
| 1993 | 402 | 0.38 | 154 | 88 | 0.73 | 64 | 314 | 0.29 | 90 | 34,803 |
| 1994 | 372 | 0.41 | 152 | 89 | 0.75 | 66 | 284 | 0.30 | 85 | 26,154 |
| 1995 | 338 | 0.44 | 148 | 89 | 0.77 | 68 | 249 | 0.32 | 80 | 17,587 |
| 1996 | 300 | 0.48 | 144 | 88 | 0.80 | 70 | 213 | 0.35 | 74 | 10,965 |
| 1997 | 265 | 0.52 | 137 | 85 | 0.83 | 70 | 180 | 0.37 | 67 | 10,412 |
| 1998 | 229 | 0.56 | 127 | 81 | 0.85 | 68 | 149 | 0.40 | 59 | 5,977 |
| 1999 | 193 | 0.59 | 115 | 74 | 0.87 | 64 | 119 | 0.42 | 51 | 3,957 |
| 2000 | 163 | 0.64 | 104 | 68 | 0.90 | 61 | 95 | 0.45 | 43 | 2709 |
| 2001 | 136 | 0.68 | 93 | 62 | 0.92 | 57 | 74 | 0.48 | 36 | 2,361 |
| 2002 | 118 | 0.74 | 87 | 59 | 0.97 | 57 | 59 | 0.50 | 30 | 2883 |
| 2003 | 102 | 0.79 | 80 | 55 | 1.01 | 56 | 46 | 0.52 | 24 | 2,378 |
| 2004 | 105 | 0.70 | 73 | 51 | 1.06 | 54 | 54 | 0.36 | 19 | 19,477 |
| 2005 | 90 | 0.75 | 67 | 47 | 1.11 | 52 | 43 | 0.36 | 15 | 500 |
| 2006 | 103 | 0.59 | 61 | 42 | 1.15 | 48 | 61 | 0.21 | 13 | 27,185 |
| 2007 | 91 | 0.59 | 54 | 37 | 1.18 | 43 | 55 | 0.20 | 11 | 4,087 |
| 2008 | 78 | 0.62 | 48 | 32 | 1.20 | 38 | 47 | 0.22 | 10 | 500 |
| 2009 | 66 | 0.64 | 43 | 27 | 1.21 | 32 | 39 | 0.26 | 10 | 300 |


|  | Proportion mature |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | $1991-1993$ | $1994-1996$ | $1997-1999$ | $2000-2002$ | $2003-2005$ | $2006-2009$ |
| 4 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 |
| 5 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 |
| 6 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 |
| 7 | 0.12 | 0.12 | 0.12 | 0.12 | 0.12 | 0.12 |
| 8 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 |
| 9 | 0.23 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 |
| 10 | 0.30 | 0.32 | 0.32 | 0.32 | 0.32 | 0.32 |
| 11 | 0.39 | 0.42 | 0.42 | 0.42 | 0.42 | 0.42 |
| 12 | 0.51 | 0.52 | 0.53 | 0.53 | 0.53 | 0.53 |
| 13 | 0.64 | 0.63 | 0.64 | 0.64 | 0.64 | 0.64 |
| 14 | 0.76 | 0.74 | 0.75 | 0.75 | 0.75 | 0.75 |
| 15 | 0.86 | 0.84 | 0.84 | 0.85 | 0.85 | 0.85 |
| 16 | 0.93 | 0.92 | 0.91 | 0.92 | 0.92 | 0.92 |
| 17 | 0.97 | 0.96 | 0.96 | 0.96 | 0.96 | 0.96 |
| 18 | 0.99 | 0.99 | 0.99 | 0.99 | 0.99 | 0.99 |
| 19 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 20 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 21 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 22 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 23 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 24 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 25 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 26 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 27 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 28 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 29 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 30 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |



Figure 7.1. Sebastes marinus in Sub-areas I and II. Total international landings 1965-2009 (in thousand tonnes)




Figure 7.2a. Illustration of the seasonality in the different Norwegian S. marinus fisheries in 2003, 2008 and 2009, also illustrating how the current regulations are working.


Figure 7.2b. Inter annual changes in the catches reported by different Norwegian S. marinus fisheries (2003-2009).


Figure 7.3. Sebastes marinus. Plot of simple mean CPUEs with 2 st. errors from the Norwegian trawl fishery, and numbers of vessel days (stippled curve) meeting the criterium of minimum $10 \%$ S. marinus in the catch per day. Upper panel shows data from the logbooks of freezer trawlers (left) and factory trawlers (right). The lower panel shows how the vessel length and use of double trawl have developed through the time series. The figure is an illustration of the data given in Table D11.


Figure 7.4a. Sebastes marinus. Abundance indices disaggregated by length for the Norwegian bottom trawl survey in the Barents Sea in winter 1986-2010 (ref. Table D12a). Top: absolute index values, bottom: relative frequencies. Horizontal lines indicates the median length in the surveyed population.


Figure 7.4b. Sebastes marinus. Abundance indices (by age) from the Norwegian bottom trawl surveys 1992-2009 in the Barents Sea (ref. Table D12b). Top: absolute index, bottom: relative frequencies. Horizontal line indicates the median age of the surveyed population.


Figure 7.5a. Sebastes marinus. Abundance indices disaggregated by length when combining the Norwegian bottom trawl surveys 1986-2008 in the Barents Sea (winter) and at Svalbard (sum$\mathrm{mer} / \mathrm{fall})$. Top: absolute index values. Bottom: relative frequencies. Horizontal line indicates the median length in the surveyed population.


Figure 7.5b. Sebastes marinus. Abundance indices disaggregated by age. Combined Norwegian bottom trawl surveys 1992-2008 in the Barents Sea (winter) and Svalbard survey (summer/fall). Top: absolute index values, bottom: relative frequencies. Horizontal line indicates median age of the surveyed population.


Figure 7.6. Sebastes marinus in Sub-areas I and II. Results from the Gadget assessment using two scientific surveys as input. The Figure shows comparison of observed and modelled survey indices (total number scaled to sum=100 during the time period) - the traditional Barents Sea February surve $y$ (top), and the coastal and fjord surve $y$ (bottom). Dots: survey indices. Phin lines: survey indices estimated by the model. Note that the 2008 and 2009 years in the coastal survey (hollow circles) have been excluded from the model tuning and the scaling.


Figure 7.7. Sebastes marinus in Sub-areas I and II. Estimates of maturity at age by Gadget. Input data have been proportions of $S$. marinus mature both at age and length as collected and classified from Norwegian commercial landings and surve ys.


Figure 7.8. Sebastes marinus in sub-areas I \& II. Unweighted average fishing mortalit y of ages 1219 as estimated by Gadget in 2010 (solid line) and in 2009 (dashed line).


Figure 7.9. Sebastes marinus in Sub-areas I and II. Estimates of abundance at age 3-6 by Gadget using two surveys as input. Gadget output provide at the 2009 AFWG are shown as dotted line. Current results are shown as plain lines.

| Total stock numbers | Total stock biomass |
| :---: | :---: |
|  |  |
| Mature stock numbers | Mature stock biomass |
| Immature stock numbers | Immature stock biomass |

Figure 7.10. Sebastes marinus in Sub-areas I and II. Stock numbers (in thousands) and biomass (in tonnes) for the total stock (3+) (upper panel), and the fishable and mature stock (middle panel), and the immature stock (lower panel), as estimated by Gadget using two surve ys as input. Gadget output provided at the 2009 AFWG are shown as dotted line. Current results are shown as plain lines.

Table D11. Sebastes marinus. Effort (vessel days) and catch per unit effort (kg per trawl hour) with 2 x st.error for Norwegian trawlers. ${ }^{1}$

|  | Freezer trawlers (32-50m) |  |  | Factory trawlers (>53m) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Number of vessel days meeting the $10 \%$ requirement | Mean CPUE per <br> year <br> (kg/hour) | $2 x$ <br> standard <br> errorof <br> the mean | Number of vessel days meeting the 10\% requirement | Mean CPUE <br> peryear <br> (kg/hour) | 2 xstandard errorof the mean |
| 1992 | 926 | 378 | 29.4 | 545 | 596 | 53.1 |
| 1993 | 743 | 374 | 34.4 | 411 | 495 | 68.9 |
| 1994 | 793 | 357 | 30.1 | 516 | 522 | 53.9 |
| 1995 | 754 | 300 | 26.7 | 343 | 323 | 35.9 |
| 1996 | 864 | 363 | 32.1 | 395 | 638 | 78.4 |
| 1997 | 972 | 331 | 31.9 | 291 | 402 | 60.3 |
| 1998 | 1303 | 230 | 17.2 | 631 | 465 | 62.1 |
| 1999 | 1054 | 224 | 18.8 | 486 | 540 | 93.1 |
| 2000 | 884 | 330 | 39.9 | 349 | 703 | 172.6 |
| 2001 | 481 | 349 | 70.5 | 421 | 753 | 118.4 |
| 2002 | 536 | 192 | 26.0 | 246 | 353 | 65.8 |
| 2003 | 276 | 136 | 21.4 | 96 | 214 | 40.7 |
| 2004 | 344 | 177 | 38.5 | 101 | 204 | 56.2 |
| 2005 | 368 | 120 | 20.2 | 160 | 160 | 24.2 |
| 2006 | 98 | 123 | 26.0 | 175 | 209 | 43.9 |
| 2007 | 147 | 167 | 29.4 | 195 | 292 | 53.5 |
| 2008 | 78 | 202 | 82.5 | 153 | 294 | 53.2 |
| $2009{ }^{2}$ | 55 | 165 | 34.4 | 104 | 331 | 129.2 |

${ }^{1}$ Only including days with more than $10 \%$ S. marinus in the catches. Only including areas with low mixing of $S$. mentella.
${ }^{2}$ Provisional figures.

Table D12a. Sebastes marinus in Sub-areas I and II. Abundance indices - on length - from the bottom trawl surveys in the Barents Sea (Division IIa) in the winter 1986-2009 (numbers in millions). The area coverage was extended from 1993.

| Year | Length group (cm) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & 5.0- \\ & 9.9 \end{aligned}$ | $\begin{aligned} & 10.0- \\ & 14.9 \end{aligned}$ | $\begin{aligned} & 15.0- \\ & 19.9 \end{aligned}$ | $\begin{aligned} & 20.0- \\ & 24.9 \end{aligned}$ | $\begin{aligned} & 25.0- \\ & 29.9 \end{aligned}$ | $\begin{aligned} & 30.0- \\ & 34.9 \end{aligned}$ | $\begin{aligned} & 35.0- \\ & 39.9 \end{aligned}$ | $\begin{aligned} & 40.0- \\ & 44.9 \end{aligned}$ | >45.0 | Total |
| 1986 | 3.0 | 11.7 | 26.4 | 34.3 | 17.7 | 21.0 | 12.8 | 4.4 | 2.6 | 133.9 |
| 1987 | 7.7 | 12.7 | 32.8 | 7.7 | 6.4 | 3.4 | 3.8 | 3.8 | 4.2 | 82.5 |
| 1988 | 1.0 | 5.6 | 5.5 | 14.2 | 12.6 | 7.3 | 5.2 | 4.1 | 3.7 | 59.2 |
| 1989 | 48.7 | 4.9 | 4.3 | 11.8 | 15.9 | 12.2 | 6.6 | 4.8 | 3.0 | 112.2 |
| 1990 | 9.2 | 5.3 | 6.5 | 9.4 | 15.5 | 14.0 | 8.0 | 4.0 | 3.4 | 75.3 |
| 1991 | 4.2 | 13.6 | 8.4 | 19.4 | 18.0 | 16.1 | 14.8 | 6.0 | 4.0 | 104.5 |
| 1992 | 1.8 | 3.9 | 7.7 | 20.6 | 19.7 | 13.7 | 10.5 | 6.6 | 5.8 | 90.3 |
| 1993 | 0.1 | 1.2 | 3.5 | 6.9 | 10.3 | 14.5 | 12.5 | 8.6 | 6.3 | 63.9 |
| 1994 | 0.7 | 6.5 | 9.3 | 11.7 | 11.5 | 19.4 | 9.1 | 4.4 | 2.8 | 75.4 |
| 1995 | 0.6 | 5.0 | 13.1 | 11.5 | 9.1 | 15.9 | 17.2 | 10.9 | 4.7 | 88.0 |
| 1996 | + | 0.7 | 3.5 | 6.4 | 9.4 | 11.7 | 16.6 | 7.9 | 3.9 | 60.1 |
| $1997{ }^{1}$ | - | 0.5 | 1.3 | 2.7 | 6.9 | 21.4 | 28.2 | 8.5 | 3.3 | 72.7 |
| $1998{ }^{1}$ | 0.1 | 3.9 | 2.0 | 7.4 | 5.8 | 25.3 | 13.2 | 7.0 | 2.3 | 67.0 |
| 1999 | 0.2 | 0.9 | 2.1 | 4.0 | 4.6 | 6.4 | 6.0 | 5.3 | 3.5 | 33.0 |
| 2000 | 0.5 | 1.1 | 1.5 | 4.2 | 4.7 | 5.0 | 3.5 | 1.8 | 1.2 | 24.0 |
| 2001 | 0.1 | 0.4 | 0.4 | 2.4 | 5.8 | 5.6 | 5.0 | 3.5 | 1.8 | 25.0 |
| 2002 | 0.1 | 1.0 | 1.9 | 1.7 | 3.7 | 4.1 | 3.3 | 3.6 | 2.5 | 22.0 |
| 2003 | 0.0 | 0.5 | 1.2 | 1.5 | 4.3 | 3.8 | 2.7 | 3.3 | 2.9 | 20.2 |
| 2004 | 0.7 | 0.2 | 0.4 | 1.0 | 2.9 | 4.4 | 5.5 | 4.0 | 3.2 | 22.3 |
| 2005 | + | 0.1 | 0.2 | 0.4 | 1.1 | 2.0 | 3.7 | 4.6 | 4.3 | 16.4 |
| 2006 | 0.0 | 0.0 | 0.0 | 0.2 | 2.5 | 5.4 | 6.1 | 4.1 | 4.2 | 22.5 |
| 2007 | 0.0 | 0.1 | 0.5 | 0.1 | 1.0 | 4.0 | 5.4 | 5.9 | 4.9 | 21.9 |
| 2008 | 1.8 | 2.6 | 0.2 | 0.2 | 0.4 | 0.7 | 1.9 | 2.5 | 4.4 | 14.8 |
| 2009 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.4 | 1.7 | 3.7 | 6.6 | 12.7 |
| 2010 | 0.4 | 2.0 | 1.2 | 0.6 | 0.1 | 0.1 | 0.8 | 1.1 | 3.9 | 10.3 |

1 - Adjusted indices to account for not covering the Russian EEZ in Subarea I

Table D12b. Sebastes marinus in Sub-areas I and II. Norwegian bottom trawl indices - on age from the annual Barents Sea survey in February 1992-2008 (numbers in thousands). The area coverage was extended from 1993 onwards.

|  | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | $\begin{array}{r} \text { Total } \\ 1-15 \\ \hline \end{array}$ | $16^{+}$ |
| 1992 | 2,295 | 4,261 | 10,760 | 2,043 | 1,474 | 13,178 | 4,230 | 6,302 | 8,251 | 3,751 | 3,865 | 3,064 | 3,568 | 67,042 | 23,300 |
| 1993 | 468 | 1,218 | 1,424 | 2,020 | 979 | 5,048 | 2,968 | 4,230 | 2,142 | 4,634 | 3,338 | 2,951 | 9,148 | 40,568 | 23,300 |
| 1994 | 2,951 | 4,485 | 2,573 | 3,801 | 8,338 | 3,254 | 1,297 | 7,231 | 6,443 | 248 | 10,192 | 6,341 | 2,612 | 59,766 | 15,600 |
| 1995 | 2,540 | 7,450 | 6,090 | 7,150 | 5,820 | 6,590 | 5,670 | 2,000 | 4,440 | 6,500 | 4,320 | 5,330 | 6,030 | 69,930 | 18,100 |
| 1996 | 310 | 1,300 | 2,340 | 3,520 | 3,660 | 8,720 | 5,650 | 3,960 | 6,590 | 5,730 | 6,230 | 4,070 | 2,950 | 55,030 | 5,100 |
| 1997 | 190 | 80 | 360 | 1,320 | 2,530 | 5,370 | 10,570 | 6,840 | 5,810 | 7,390 | 8,790 | 9,740 | 1,980 | 60,980 | 11,700 |
| 1998 | 2,380 | 1,930 | 850 | 660 | 1,140 | 7,090 | 6,124 | 4,962 | 4,091 | 5,190 | 8,790 | 2,730 | 2,560 | 48,487 | 18,500 |
| 1999 | 737 | 916 | 1,246 | 3,469 | 1,650 | 1,826 | 1,679 | 3,084 | 2,371 | 2,953 | 3,837 | 2,132 | 1,979 | 27879 | 5,100 |
| 2000 | 490 | 720 | 900 | 1,310 | 1,800 | 2,440 | 2,020 | 2,710 | 2,090 | 940 | 1,440 | 2,940 | 430 | 20,230 | 3,800 |
| 2001 | 320 | 170 | 190 | 940 | 1,360 | 2,220 | 3,110 | 2,400 | 2,690 | 2,230 | 2,180 | 1,200 | 1,370 | 20380 | 4,600 |
| 2002 | 130 | 910 | 902 | 1,590 | 544 | 1,546 | 2,153 | 1,822 | 1,900 | 2,220 | 1,073 | 1,294 | 1,730 | 17814 | 4,200 |
| 2003 | 220 | 250 | 590 | 1,080 | 680 | 1,020 | 2,910 | 1,180 | 2,250 | 1,370 | 1,530 | 840 | 1,310 | 15,230 | 5,000 |
| 2004 | 780 | 100 | 100 | 90 | 240 | 540 | 1,130 | 1,260 | 1,590 | 1,740 | 1,490 | 2,570 | 1,890 | 13,520 | 8,800 |
| 2005 | 39 | 85 | 107 | 110 | 321 | 524 | 669 | 497 | 697 | 820 | 1,517 | 1,905 | 1,653 | 8,944 | 7,652 |
| 2006 | 0 | 0 | 0 | 24 | 52 | 1,011 | 1,641 | 1,999 | 2,246 | 1,578 | 1,550 | 3,487 | 1,444 | 15,030 | 7,666 |
| 2007 | 58 | 202 | 248 | 50 | 51 | 185 | 422 | 582 | 592 | 1,747 | 1,030 | 1,127 | 1,359 | 7,652 | 14,248 |
| 2008 | 2637 | 0 | 0 | 0 | 203 | 72 | 175 | 272 | 476 | 369 | 553 | 850 | 700 | 6,306 | 6,543 |
| 2009 | 0 | 0 | 0 | 0 | 85 | 0 | 14 | 77 | 192 | 358 | 1,146 | 532 | 737 | 3,141 | 9,539 |

${ }^{1} 16+$ group is considered in the calculation since 2005. Values prior to this date were derived by subtracting the sum of abundance in groups 1-15 to the total abundance, available in Table D12a.

Table D13a. Sebastes marinus in Subarea I and II. Abundance indices - on length - from the bottom trawl survey in the Svalbard area (Division IIb) in summer/fall 1985-2008 (numbers in thousands).

|  | Length group (cm) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | $\begin{array}{r} 5.0- \\ 9.9 \end{array}$ | $\begin{array}{r} 10.0- \\ 14.9 \end{array}$ | $\begin{array}{r} 15.0- \\ 19.9 \end{array}$ | $\begin{array}{r} 20.0- \\ 24.9 \end{array}$ | $\begin{array}{r} 25.0- \\ 29.9 \end{array}$ | $\begin{array}{r} 30.0- \\ 34.9 \end{array}$ | $\begin{array}{r} 35.0- \\ 39.9 \end{array}$ | $\begin{array}{r} 40.0- \\ 44.9 \end{array}$ | >45.0 | Total |
| $1985{ }^{1}$ | - | 1,307 | 795 | 1,728 | 2,273 | 1,417 | 311 | 142 | 194 | 8,325 |
| $1986{ }^{1}$ | 200 | 2,961 | 1,768 | 547 | 643 | 1,520 | 639 | 467 | 196 | 8,941 |
| $1987{ }^{1}$ | 100 | 1,343 | 1,964 | 1,185 | 1,367 | 652 | 352 | 29 | 44 | 7,060 |
| $1988{ }^{1}$ | 500 | 1,001 | 1,953 | 1,609 | 684 | 358 | 158 | 68 | 95 | 6,450 |
| 1989 | 200 | 1,629 | 2,963 | 2,374 | 1,320 | 846 | 337 | 323 | 104 | 10,100 |
| 1990 | 1,700 | 3,886 | 4,478 | 4,047 | 2,972 | 1,509 | 365 | 140 | 122 | 19,185 |
| 1991 | 100 | 5,371 | 5,821 | 9,171 | 8,523 | 4,499 | 1,531 | 982 | 395 | 36,420 |
| 1992 | 1,700 | 10,228 | 8,858 | 5,330 | 13,960 | 12,720 | 4,547 | 494 | 346 | 58,172 |
| 1993 | 200 | 10,160 | 9,078 | 5,855 | 7,071 | 4,327 | 2,088 | 1,552 | 948 | 41,284 |
| 1994 | 100 | 3,340 | 5,883 | 4,185 | 3,922 | 3,315 | 1,021 | 845 | 423 | 22,985 |
| 1995 | 470 | 2,000 | 9,100 | 5,070 | 3,060 | 2,400 | 1,040 | 920 | 780 | 24,840 |
| 1996 | 80 | 130 | 1,260 | 2,480 | 1,030 | 480 | 550 | 990 | 400 | 7,400 |
| 1997 | 0 | 810 | 1,980 | 5,470 | 5,560 | 2,340 | 590 | 190 | 450 | 17,430 |
| 1998 | 180 | 2,698 | 1,741 | 4,620 | 4,053 | 1,761 | 535 | 545 | 241 | 16,403 |
| 1999 | 0 | 794 | 7,057 | 3,698 | 4,563 | 2,449 | 467 | 619 | 369 | 20,017 |
| 2000 | 40 | 360 | 1,240 | 1,390 | 2,010 | 760 | 400 | 160 | 390 | 6,750 |
| 2001 | 10 | 110 | 790 | 1,470 | 3,710 | 4,600 | 1,880 | 680 | 370 | 13,660 |
| 2002 | 0 | 0 | 64 | 415 | 459 | 880 | 620 | 565 | 519 | 3,522 |
| 2003 | 90 | 90 | 108 | 83 | 525 | 565 | 447 | 760 | 769 | 3,437 |
| 2004 | 0 | 0 | 10 | 50 | 650 | 740 | 670 | 430 | 190 | 2,740 |
| 2005 | 0 | 45 | 0 | 30 | 315 | 384 | 307 | 159 | 274 | 1,513 |
| 2006 | 0 | 0 | 70 | 64 | 167 | 376 | 473 | 735 | 1,514 | 3,398 |
| 2007 | 0 | 32 | 58 | 1,003 | 1,049 | 3,875 | 4,656 | 811 | 1,267 | 12,751 |
| 2008 | 7,009 | 3,573 | 175 | 21 | 42 | 142 | 475 | 162 | 529 | 12,130 |
| 2009 | 227 | 1,476 | 114 | 114 | 0 | 0 | 185 | 213 | 193 | 2,522 |

1 - Old trawl equipment (bobbins gear and 80 meter sweep length)

Table D13b. Sebastes m arinus in Sub-areas I and II. Norwegian bottom trawl survey indices - on age - in the Svalbard area (Division IIb) in summer/fall 1992-2008 (numbers in thousands).

|  | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | Total |
| 1992 | 284 | 12,378 | 5,576 | 2,279 | 371 | 2,064 | 3,687 | 5,704 | 9,215 | 6,413 | 1,454 | 1,387 | 696 | 22 | 51,530 |
| 1993 | 32 | 10,704 | 5,710 | 5,142 | 1,855 | 1,052 | 1,314 | 3,520 | 2,847 | 2,757 | 2,074 | 1,245 | 844 | 119 | 39,215 |
| 1994 | 429 | 1,150 | 3,418 | 2,393 | 1,723 | 1,106 | 1,714 | 1,256 | 1,938 | 1,596 | 2,039 | 484 | 550 | 319 | 20,155 |
| 1995 | 600 | 1,600 | 6,400 | 5,100 | 1,800 | 2,200 | 1,800 | 700 | 700 | 400 | 700 | 500 | 400 | 500 | 23,400 |
| 1996 | 40 | 110 | + | 560 | 1,050 | 940 | 930 | 400 | 1,050 | 280 | 320 | 590 | 160 | 70 | 6,500 |
| 1997 | 320 | 490 | + | 480 | 1,500 | 6,950 | 2,720 | 1,680 | 800 | 1,310 | 550 | 30 | + | 120 | 16,950 |
| 1998 | 210 | 1,817 | 881 | 202 | 1,555 | 2,187 | 4,551 | 1,913 | 1,010 | 797 | 49 | 264 | 73 | 187 | 15,696 |
| 1999 | 0 | 760 | 2,893 | 1,339 | 3,534 | 1,037 | 3,905 | 2,603 | 762 | 1,663 | 481 | 361 | 258 | 152 | 19,748 |
| 2000 | 40 | 20 | 400 | 350 | 840 | 480 | 730 | 1,670 | 620 | 340 | 510 | 100 | 80 | 70 | 6,250 |
| 2001 | 0 | 40 | 50 | 450 | 330 | 790 | 1,760 | 1,970 | 3,300 | 1,200 | 1,810 | 150 | 660 | 430 | 12,940 |
| 2002 | 0 | 0 | + | + | 65 | 160 | 204 | 326 | 364 | 614 | 442 | 328 | 15 | 0 | 2,518 |
| 2003 | 30 | 30 | 30 | + | 108 | + | 219 | 263 | 126 | 259 | 306 | 199 | 248 | 411 | 2,229 |
| 2004 | 0 | 0 | 0 | + | + | 20 | 360 | 120 | 430 | 160 | 410 | 360 | 370 | 200 | 2,430 |
| 2005 | 0 | 45 | 0 | 0 | 0 | 30 | 48 | 228 | 138 | 187 | 194 | 93 | 105 | 109 | 1,177 |
| 2006 | 0 | 0 | 23 | 23 | 23 | 21 | 22 | 21 | 84 | 0 | 84 | 279 | 194 | 376 | 1,148 |
| 2007 | 0 | 33 | 19 | 19 | 19 | 764 | 764 | 525 | 0 | 0 | 21 | 1,927 | 1,927 | 1,683 | 7,702 |
| 2008 | 10583 | 44 | 88 | 44 | 11 | 11 | 0 | 42 | 88 | 13 | 13 | 118 | 63 | 174 | 11,292 |

Table D14. Sebastes marinus in Sub-area I and II. Mean catch rates (N/nm2) of Sebastes marinus from Norwegian Coastal Surveys (Division IIa) in 1995-2009 within 100-350 m depth. Catch rates for the total area.

| Length range (cm) | H | à | $\stackrel{\underset{1}{\top}}{\underset{\sim}{6}}$ | $\begin{aligned} & \stackrel{\rightharpoonup}{1} \\ & \stackrel{1}{2} \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { N } \\ & \text { N } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { N̦ } \\ & \text { L } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { W్ } \\ & \text { oे } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { oे } \\ & \text { مि } \\ & \hline \end{aligned}$ | $\begin{aligned} & 7 \\ & \hline \end{aligned}$ | $\begin{aligned} & 9 \\ & 4 \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { Ĥ } \\ & \text { O } \\ & \hline 1 \end{aligned}$ | $\begin{aligned} & \text { ir } \\ & \text { L? } \\ & 10 \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { H } \\ & \text { O } \\ & \text { O} \end{aligned}$ |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 0 | 41 | 118 | 59 | 54 | 38 | 69 | 214 | 157 | 21 | 2 | 1 | 0 |  |  |  |  |  |
| 1996 | 0 | 34 | 87 | 124 | 151 | 67 | 210 | 415 | 209 | 64 | 0 | 0 | 0 |  |  |  |  |  |
| 1997 | 0 | 4 | 9 | 12 | 64 | 112 | 96 | 178 | 190 | 45 | 2 | 1 | 0 |  |  |  |  |  |
| 1998 | 0 | 0 | 0 | 4 | 12 | 16 | 17 | 110 | 96 | 18 | 3 | 0 | 0 |  |  |  |  |  |
| 1999 | 0 | 0 | 19 | 242 | 160 | 34 | 43 | 151 | 117 | 15 | 4 | 2 | 0 |  |  |  |  |  |
| 2000 | 0 | 0 | 2 | 13 | 7 | 10 | 30 | 160 | 155 | 30 | 4 | 0 | 0 |  |  |  |  |  |
| 2001 | 0 | 0 | 2 | 11 | 14 | 22 | 15 | 83 | 160 | 30 | 2 | 0 | 0 |  |  |  |  |  |
| 2002 | 0 | 0 | 0 | 0 | 2 | 6 | 29 | 259 | 213 | 26 | 4 | 1 | 0 |  |  |  |  |  |
| 2003 | 0 | 0 | 6 | 10 | 43 | 66 | 49 | 219 | 225 | 55 | 6 | 1 | 2 | 123 | 160 | 1367 | 1053 | 43574 |
| 2004 | 0 | 1 | 3 | 6 | 21 | 66 | 35 | 351 | 552 | 42 | 3 | 1 | 0 | 104 | 130 | 1290 | 950 | 43574 |
| 2005 | 0 | 1 | 5 | 5 | 30 | 46 | 48 | 190 | 171 | 37 | 1 | 0 | 0 | 99 | 132 | 833 | 780 | 43574 |
| 2006 | 0 | 0 | 3 | 0 | 2 | 3 | 30 | 145 | 256 | 66 | 9 | 0 | 0 | 112 | 112 | 771 | 680 | 43574 |
| 2007 | 0 | 0 | 0 | 0 | 4 | 7 | 17 | 129 | 177 | 29 | 1 | 0 | 0 | 131 | 140 | 637 | 637 | 43574 |
| 2008 | 0 | 4 | 5 | 1 | 4 | 5 | 17 | 363 | 490 | 99 | 12 | 2 | 0 | 110 | 140 | 1156 | 850 | 43574 |
| 2009 | 0 | 0 | 8 | 3 | 10 | 19 | 45 | 808 | 945 | 109 | 14 | 1 | 0 | 109 | 127 | 2945 | 581 | 43574 |



Figure D 1. Overview of the Norwegian biological samples from the commercial fisheries for $S$. marinus in 2008 representing more than $80 \%$ of the catches and which the input data to the Gadget model are based upon. The colours denote which sampling platform has been used: port sampling (black), Reference fleet (light blue), inspectors/observers (dark blue, green), winter (yellow) and summer (pink) scientific surve ys. The crosses show the catch in tonnes for the different seasons, areas and gear.

## 8 Greenland halibut in subareas I and II

An update assessment is presented for this stock. This should be regarded as an exploratory run and just used to view trends in the stock. The work on the age reading problems are continued, but we still need time before a thorough benchmark assessment can be carried out. General information about this stock is located in the Quality Handbook.

### 8.1 Status of the fisheries

### 8.1.1 Landings prior to 2010 (Tables 8.1-8.5, E10)

Nominal catches by country for Subareas I and II combined are presented in Table 8.1. Tables 8.2-8.4 give the catches for Subarea I and Divisions IIa and Ilb separately, and landings separated by gear type are presented in Table 8.5. For most countries the catches listed in the tables are similar to those officially reported to ICES. Some of the values in the tables vary slightly from the official statistics, and represents those presented to the W orking Group by the members.

The preliminary estimate of the total catch for 2009 is $12,207 \mathrm{t}$. This is about $6 \%$ less than the projected catch for 2009 estimated by the Working Group during its 2009 meeting ( $13,000 \mathrm{t}$ ). It is also the lowest catch since 1999. The difference between projected catch and preliminary estimate of total catch for 2009 is mainly due to Russian landings being low er than projected.

Some fishing for Greenland halibut has taken place in the northern part of Division IVa during the past 20-30 years, varying between a few tonnes and up to 2,500 $t$ in 1999. Since 2005 this catch has been mostly below 100 t , and in 2009 it was 134 t taken mostly by UK (Table E10). This fishery is in another management area, and is not restricted by any TAC regulations. Although there is a continuous distribution of this species from the southern part of Division IIa along the continental slope towards the Shetland area, little is known about the stock structure and the catch taken from this area has therefore not been added to the catch from Subareas I and II.

Around Jan Mayen, small catches of Greenland halibut have been taken in some years. 21 t were reported from this area in 2006, whereas in 2007-2009 no catches were reported. Jan Mayen is within Subarea Ila, but little is known about the relationship with the stock assessed by the Arctic Fisheries Working Group. Catches from this area have therefore not been included in the catches given for Subarea II.

### 8.1.2 ICES advice applicable to 2009 and 2010

The advice from ICES for 2009 was as follows:
Exploitation boundaries in relation to precautionary limits: The stock has remained at a relatively low size in the last 25 years at catch levels of 1500025000 t . In order to increase the SSB, catches should be kept well below that range. Catches for 2008 should be below 13000 t as advised since 2003; this is the level below which SSB has increased in the past.

Exploitation boundaries in relation to high long-term yield, low risk of depletion of production potential and considering ecosystem effects: There is no estimate of highyield reference points.

The advice for ICES for 2010 was as follows:

Exploitation boundaries in relation to precautionary limits: The stock has remained at a relatively low size in the last 25 years at catch levels of 1500025000 t . In order to increase the SSB, catches should be kept well below that range. Catches for 2009 should bebelow 13000 t as advised since 2003; this is the level below which SSB has increased in the past.

Exploitation boundaries in relation to high long-term yield, low risk of depletion of production potential and considering ecosystem effects: There is no estimate of highyield reference points.

### 8.1.3 Management applicable in 2009 and 2010

Target Greenland halibut fishery was forbidden since 1992 until 2009. Management of Greenland halibut is by bycatch regulations and a limited coastal Norwegian fishery using longline and gillnet. From 2001 the bycatch regulations in each haul was not to exceed $12 \%$ in each haul and $7 \%$ of the landed catch. From early 2004 the Norwegian Department of Fisheries decided that for Norw egian vessels in the NEEZ allowable bycatch at any time on board and by landing should not exceed $7 \%$. In addition, the annual catch for each trawler are not allowed to exceed $4 \%$ of the sum of the vessels quota on cod, haddock and saithe, and limited by a maximum annual catch of 40 t pr. vessel.

The Norwegian conventional fleet, vessels smaller than 28 m , are allowed to conduct a limited target fishery with longlines and gillnets in a limited area in approximately one month each year. For these vessels the TAC is set to 10,12 and 14 t , dependent of size of the vessel. This fishery is supposed to keep the total catch at a level which these vessels landed historically (ca. 2,500 t).

The $30^{\text {th }}$ Session of the joint Russian-Norwegian Fisheries Commission (JRNFC) in 2001 stated that both the Russian and the Norwegian party could catch up to $1,500 \mathrm{t}$ of Greenland halibut for research and surveillance purposes in 2002. This research quota was increased in the commission meeting the year after to $3,000 \mathrm{t}$ for each party, and stayed at this level until 2005. The JRNFC then increased the quota to $4,500 \mathrm{t}$ for each party in 2006, and $4,900 \mathrm{t}$ for each party in 2007. During the $36^{\text {th }}$ Session of the JRNFC it was decided to decrease quotas for 2008 to $4,000 \mathrm{t}$ for each party. The $3^{\text {th }}$ JRNFC's Session kept the research quotas for 2009 at the same level.

The $38^{\text {th }}$ JRNFC's Session in 2009 decided to cancel the ban against target Greenland halibut fishery and established the TAC at $15,000 \mathrm{t}$ within next three years (20102012). The TAC was allocated between Norway, Russia and other countries with shares 51,45 and $4 \%$ respectively.

### 8.1.4 Expected landings in 2010

Due to new regulation measures established in 2009 for 2010-1012, the total Greenland halibut catch in the Barents Sea and adjacent waters (ICES Subarea I and Divisions IIa and $\Pi 1 b$ ) in 2010 is expected to be about $15,000 \mathrm{t}$. Discards is not regarded as a problem, but it is believed that there may be additional landings that are not reported. The catches from Division IVa are expected to be maintained at a low level (below 100 t ).

### 8.2 Status of research

### 8.2.1 Survey results (Tables A14, E1-E8)

Over the last several years the W orking Group has been concerned about trends in catchability within individual surveys used for tuning of the XSA. The trends were seen for younger ages of year classes in the late 80's and early 90's that were initially estimated very low in abundance. With increasing age these year classes were estimated much closer to the mean abundance. In previous meetings the Working Group therefore increased the lower age used in tuning to five years in order to reduce the problem. This only partly solved the problem, and in all subsequent assessments estimated recruitment of the last 2-3 years increased from one year to the next.

Most of the surveys considered by the Working Group covered either the adult population in the slope area or juvenile distribution in northern areas. The problem of underestimation of recruitment in the last few years included in the analyses was attributed to shortcomings in survey coverage. At previous meetings, the Working Group had noted the need for annual surveys that sample most of the population within a short period of time. Prior to the 2002 Working Group meeting, effort was therefore made to combine some of these surveys into a new total index. The new index was termed the Norwegian Combined Survey Index and was established back to 1996, the first year with survey coverage northeast of Svalbard. It includes bottom trawls from the Norwegian bottom trawl survey in August in the Barents Sea and Svalbard (Tables E1 and E2), the Norwegian Greenland halibut survey in August along the continental slope (Table E3), and the Norwegian bottom trawl survey in August-September north and east of Svalbard (Table E4). With exception of the Norwegian Greenland halibut survey, all these surveys were from 2004 conducted as one major joint survey between Norway and Russia. Prior to the meeting in 2003, work was done to evaluate the combination of these survey series into one index, and this was reported to the Working Group (Pennington, WD 5\#2003). Based on these results it was decided to use the combined index in the assessment. Although representing a larger part of the stock, the new combined survey indices were not successful in establishing consistency in the relative size of year classes at age. Future inclusion of northern parts of the Russian zone may improve the index. The W orking Group has later advised that further work should be done to improve the combined index with regards to pooling different surveys using different gears.

Also in the Russian bottom trawl surveys in October-December (Table E6) it has been difficult to identify year classes that appear consistently either strong or weak across ages. In previous Working Group reports this survey series was the one with the clearest and strongest trends in catchability with age in the XSA calibrations. These surveys are important since they usually cover large parts of the total known distribution of the Greenland halibut within 100-900 m depth. However, it has been considered imprudent to use the 2002 and 2003 data from this survey series. During the 2002 survey, no observations were available from the Exclusive Economic Zone of Norway (NEEZ). In 2003, observations on the main spawning grounds were conducted three weeks later than usual because access to NEEZ was obtained too late. The number of trawl stations was also insufficient due to the same reason.

The Norwegian CPUE survey (Table E9) was stopped from 2005. This was one of the tuning fleets, but an evaluation of this survey revealed a lot of inconsistencies in the
series. Since 2006, none of the age structured tables of the Norwegian surveys have been updated due to change in age reading procedure.

Thejoint Russian-Norw egian research program on Greenland halibut had finished in 2009 and will eventually contribute by increasing the understanding of the occuring processes. One of the main objectives of the program was to clarify the migration dynamics of the stock, including vertical distribution and relations with Greenland halibut in other areas. The results may improve both biological sampling and the subsequent assessments. The project has developed a new age reading procedure which has been used in Norway since 2006. This will eventually end up in total revision of the input data to the assessment.

During the last ten years there was a slowly increasing trend in biomass estimates both from the Norwegian CPUE survey (ended in 2006), the Norw egian Combined Index and the Russian Index (Figure 8.4). However, the biomass indices of mature females from different surveys showed opposite trends in last years (Figure 8.5).

The Spanish bottom trawl survey from 1997 to 2008 (Table E7) showed an increase of Greenland halibut abundance and biomass in the Svalbard-Bear Island area from 2002 after three years with a declining trend.

Abundance indices of 0-group Greenland halibut are shown in Table 1.1. The increase in 0-group abundance after 1996 seems to have stopped, and the 2007-2009 indices were very low. It should be noted that the Ecosystem survey is not optimal for surveying 0 -group Greenland halibut.

### 8.2.2 Commercial catch-per-unit-effort (Table 8.6 and E9)

The CPUE from the experimental fishery was found to be considerably higher than in the traditional fishery and has exhibited an increasing trend from 1992-1996. After 1996 the Norwegian CPUE series has varied between 1200 and $1800 \mathrm{~kg} / \mathrm{h}$ with the highest value in 2005 (Table E9). The Russian experimental CPUE series shows an increasing trend since 1997, and this series shows the highest value in 2003. A significant decline was observed in 2004-2008 (Table 8.6) and in 2009 indices jump up again. The Norwegian CPUE survey was terminated in 2006.

### 8.2.3 Age readings

Based on scientific presentiment that the species is more slow growing and vulnerable than the previous age readings suggest, the Norwegian age reading were changed in 2006. The new Norwegian age readings are- not comparable with older data or the Russian age readings. Age reading problems are addressed in the joint research program, and this will lead to revised age structure in the input data in the future. There are some uncertainties to when these revised age readings can be used in the assessment. It is of outmost importance that scientists that are engaged in age reading on Greenland halibut from all involved member states will participate in the ICES ageing w orkshop on Greenland halibut (WKARGH) in Vigo 14-17 February 2011. The w orkshop will be a milestone towards age reading method for this species that is accepted and agreed on.

In 2007-2010, Russian age-length keys were used on the total catch matrix and the Russian survey was the only tuning fleet being updated. The two Norwegian surveys were used as before as tuning series until 2005.

### 8.3 Data used in the assess ment

Based on the arguments in Section 8.2.1 the Working Group also this year considers the survey indices for ages below age 5 not appropriate for inclusion in the tuning data. Consequently, a standard XSA was run for age 5 and above.

### 8.3.1 Catch-at-age (Table 8.7)

The catch-at-age data for 2008 were updated using revised catch figures. Catch-atage data for 2006-2009 were available only from the Russian fisheries. The Russian age-lenght keys were used to allocate catches from the other countries by age groups. Total international catch-at-age is given in Table 8.7. Greenland halibut are usually caught in the range of $3-16$ years old, but the catch is mainly dominated by ages 510. Generally, fish older than age 10 comprise a very low proportion of the catches.

### 8.3.2 Weight-at-age (Table 8.8)

For the years 1964-1969 separate weight-at-age data were used for the Norwegian and the Russian catches. Both data sets were mean values for the period and were combined as a weighted average for each year. A constant set of weight-at-age data was used for the total catches in the years 1970-1978. For subsequent years annual estimates were used. The Russian weight-at-age data was used in the catch in 20062009 (Table 8.8). The weight-at-age in the stock was set equal to the weight-at-age in the catch for all years.

### 8.3.3 Natural mortality

Natural mortality of Greenland halibut was set to 0.15 for all ages and years. This is the same assumption as was used in previous years.

### 8.3.4 Maturity-at-age (Tables 8.9)

Annual ogives were derived to estimate the spawning stock biomass based on females only using Russian survey data for the years 1984-2008, except for the year 1991. An average ogive computed for 1984-1987 was applied to 1964-1983. The average of 1990 and 1992 was used to represent the maturity ogive for 1991. For 19842002 and 2004-2009 a three-year running average was applied. In previous assessments a similar procedure using the same data set was implemented but was based on sexes combined. The ogive for 2003 was rejected due to the problems with the Russian survey mentioned above (Section 8.2.1) and the data used was the mean value for 2002 and 2004.

### 8.3.5 Tuning data

The XSA was run with the same tuning series as used in last year's assessment:
Fleet 4: Experimental commercial fishery CPUE from 1992-2005 for ages 5-14.
Fleet 7: Russian trawl survey from 1992-2009 for ages 5-14. The 2002 and 2003 data was not included in this series due to the problems mentioned in section 8.2.1

Fleet 8: Norwegian Combined Survey from 1996-2005 for ages 5-15.
The software XXSA.exe were used.

### 8.4 Recruitment indices (Tables A14, E1-E9)

In addition to the indices mentioned in Section 8.3.5, all surveys in Section 8.2.1 may provide information on recruitment. However, because the dynamics of migration and distribution patterns are not well understood for this stock, it is not known which age should be used for a reliable recruitment estimate. As outlined in previous Working Group reports there is no longer evidence for a major recruitment failure in the 1990's. Nevertheless, the relative size of the individual year classes is still poorly estimated, especially at ages below 5 years.

### 8.5 Methods used in the assessment

### 8.5.1 VPA and tuning (Figure 8.1, Tables 8.7-8.10)

The Extended Survivors Analysis (XSA) was used to tune the VPA to the fleets as mentioned in Section 8.3.5. The analyses used survivor estimates shrunk towards the mean of the final 2 years and 5 ages and the standard error of the mean to which the estimates were shrunk was set to 0.5 . The catchability was considered to be independent of stock size for all ages and independent of age for ages 10 and older. These are the same settings as used in last years assessment.

Input data and diagnostics of the final XSA run are given in Tables 8.7-8.10 and log catchability residuals for the three fleets used in the tuning are shown in Figure 8.1.

### 8.6 Results of the Assessment

The diagnostics of the assessment indicate that it is generally unbiased, and describes the trend in stock development reasonably well. The survivor estimates for 2008 for most of the important year classes are determined primarily from the tuning fleet data and in most instances each tuning fleet contributes significantly to the determinations with little effect from inclusion of F shrinkage means in the tuning process. Nevertheless, the assessment diagnostics also indicated substantial uncertainties in absolute values of the survivor estimates determined by the analysis shown by instances of very high residuals, large S.E. $(\log q$ )'s and low R2's in the regression statistics for certain fleets and ages.

### 8.6.1 Results of the VPA (Figure 8.2, Tables 8.11-8.15)

The fishing mortality (F) matrix indicates that historically Greenland halibut were fully recruited to the fishery at approximately age 6-7. Since 1991 the age of full recruitment appears closer to age 10 (Table 8.11). This is likely due to a substantial proportional reduction in trawler effort since 1991 combined with reduced catchability of some year classes in the fishing areas. Trawlers catch more young fish compared to gillnetters and longliners. Nevertheless, F on ages 6-10 continues to represent the average fishing mortality on the major age groups prosecuted by the fishery.

Until 1976 the female spawning stock varied between 60,000 and $140,000 \mathrm{t}$, then it was relatively stable at around 40,000 t until the mid 1980's after which it declined markedly. It reached an all time low of $14,800 \mathrm{t}$ by 1995-96 but has been increasing since then to an estimate of $59,000 \mathrm{t}$ by 2004, which is the highest value estimated since 1976 and higher than the long-term average for the whole period 1964-2009. The female spawning stock has decreased in 2005-2007 and increased again in 2009. The total stock decreased from $312,000 \mathrm{t}$ in 1970 to the historical minimum at $46,000 \mathrm{t}$ in 1992 and then shows a positive trend with the highest estimates at about $167,000 \mathrm{t}$
in 2009. The maturity ogives used has shown a very variable maturity by age in the recent years and this affects the SSB.
Prior to the reduction in the early 1990's the fishing mortality had increased continuously for more than a decade and peaked in 1991 at 0.65 . After the reduction the fishing mortality has averaged around 0.25 . The high catch in 1999 resulted in an increase in fishing mortality to 0.34 but has since then declined to $0.14-0.15$ by 2002 and 2003, the lowest value estimated for the last 20 years. Due to the increased catch in 2004-2006 the fishing mortality again slightly raised (0.17-0.18) but remained lower than average. For the 2009 Fbar was estimated at 0.08 - the same as in 2008 which is lowest level in history. It was conditioned by stock growth and significant reducing of total catch.

Recruitment-at-age 5 in this year assessment shows the huge increase from 2007 to 2009. The 2009 level at 52 millions specimens occured twice higher than long-term average (table 8.15).

### 8.6.2 Biological reference points

Given the continuing levels of uncertainty in the current assessment no further attempts were made to develop reference points for this stock.

### 8.6.3 Catch options for 2011

Given the uncertainty around the absolute values of population size at age no catch options are provided.

### 8.7 Comparison of this years assessment with last years assessment

Compared to last year assessment stock size for 2008 has sharply increased while SSB has been little bit reduced, fishing mortality remained at nearly the same level.

|  | TOTAL STOCK (5+) BY | SSB BY | F6-10 iN 2009 | F6-10 in 2008 |
| :--- | :--- | :--- | :--- | :--- |
|  | 1 JANUARY 2009 | 1 JANUARY 2009 |  |  |
| WG 2009 | 127097 | 42255 | $0.13^{*}$ | 0.09 |
| WG 2010 | 160481 | 41526 | 0.08 | 0.08 |

*prediction

### 8.8 Comments to the assessment (Figures 8.3-8.6)

The assessment was classified as an update assessment. The current assessment was using the same catch matrix, surveys series and settings as in the previous year with updated data for 2008 and new data for 2009. Fishing mortalities tend to be overestimated while SSB tends to be underestimated in the assessment year as illustrated by the retrospective plots in Figure 8.3.

The assessment is considered to be still uncertain due to the age-reading and survey data quality problems. Nevertheless the assessment may be accepted as indicative for stock trends. Although many aspects of the assessment remain uncertain, most fishery independent indices of stock size indicate positive trends in recent years. The biomass indices from the two Norwegian survey series seem to level out in later years. (Figure 8.4).

The main result from the assessment is that the total stock has an increasing trend since 1992 and this is also seen in the SSB from 1995 to 2004. In 2004-2008 the SSB show a decreasing signal, whereas it has a significant increase in 2009. The estimate
of the SSB is based on maturity ogives from the Russian survey. Other sources indicates no decreasing trend in the maturity of Greenland halibut in recent years. Biomass indices of mature females from the Norwegian survey in the slope area (main adult area) had opposite trends untill 2008, but showes increase in 2009 (Figure 8.5). However, estimates from the Russian December survey show decrease in mature female biomass betw een 2008 and 2009.

A WD was presented to the meeting where the XSA diagnostics were scrutinized (Hallfredsson WD 18). Based on this scrutiny XSA runs were conducted where canges were made in defination of plus group and the age from where catchability is considered constand in the analysis. The results showd that the current XSA is senistive to these changes regarding especially SSB estimates and to some extend the trends in SSB (figure 8.6). The sensitivity to these relatively minor modifications in model assumptions one more time confirms the nessesity of carefulness in settings selection and possible shortcomings in input data.

Also the presentation was distributed on a GIS based assessment for Greenland halibut (Bulatov and Moiseenko). According to them in 1998-2009 the average value of fishablebiomass of Greenland halibut exceeded 347 thou.t. This method has been presented to AFWG previously for other stocks, e.g. NEA cod (AFWG report 2007) and some problems were identified:

- First, the use of catch rates from commercial fishing vessels to obtain swept area estimates representative for larger areas violates the condition that such measures of density have to be based on random samples. Obviously, fishing vessels do not fish at random, but use former experience and various fish-finding tools to seek up the densest concentrations before setting the trawls. Consequently, the catch rates obtained are only representative for the area covered by the trawl during the haul.
- Second, the method uses a constant trawl catchability factor for all length groups, trawl types, seasons etc., which is highly questionable.
- Third, a width of trawling equal to the wingspread of the trawl is used, not taking into consideration the herding effect of trawl wires, trawl doors, and sweeps.

The same problems are identified for the present analysis on Greenland halibut. The first point is a fundamental problem, which does not allow for this method to be used for absolute abundance estimation. Additionally one can question if data from the fisheries on trawl geometry and trawling distance, as required for the analysis, are sufficiently reliable for absolute abundance estimates. Consequently, this method has potential for use as an index of relative abundance that can be used as an additional tuning series for a VPA, but cannot be used as absolute abundance estimates.

The Working group have stated in several previous reports that catches above the mean in the period 1992-2003 (ca. 13,000 t) reduces the stocks ability to rebuild. The quite low catch in 2008 and 2009 will most likely lead to further growth of the both total and spawning stock size.

Average catch during the period 2004-2009 have consisted ca. 16,000 t.

### 8.9 Response to ACOM technical minutes

ACOM technical minutes are not commented on because the 2010 advice should be "same as previous year" and the report will not be reviewed. There were few
technical comments and most of them relate to ageing, and cannot be solved until there is a consensus on age reading methods. A workshop on age reading of Greenland halibut will be held in February 2011.

Table 8.1. GREENLAND HALIBUT in Sub-areas I and II. Nominal catch (t) by countries (Subarea I, Divisions IIa and IIb combined) as officially reported to ICES.

${ }^{1}$ Provisional figures.
2 Working Group figures.
${ }^{3}$ As reported to Norwegian authorities.
${ }^{4}$ USSR prior to 1991.

TABLE 8.2. GREENLAND HALIBUT in Sub-areas I and II. Nominal catch (t) by countries in Subarea I as officially reported to ICES.

| Year | Estonia | Faroe <br> Isl ands | Fed. Rep. <br> Germany | France | Green- <br> land | Ice- <br> land |  | Norway | Poland | Portugal |  | Russia ${ }^{4}$ |  | $\begin{gathered} \hline \text { UK } \\ (\mathrm{E} \& \mathrm{~W}) \end{gathered}$ | $\begin{gathered} \hline \text { UK } \\ \text { (Scot.) } \end{gathered}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | - | - | - |  | - - | - | - | 593 |  | - | - | 81 | - | 17 | - | 691 |
| 1985 | - | - | - | - | - - | - | - | 602 |  | - | - | 122 | - | 1 | - | 725 |
| 1986 | - | - | 1 |  | - - | - | - | 557 |  | - | - | 615 | - | 5 | 1 | 1,179 |
| 1987 | - | - | 2 |  | - - | - | - | 984 |  | - | - | 259 | - | 10 | + | 1,255 |
| 1988 | - | 9 | 4 |  | - - | - | - | 978 |  | - | - | 420 | - | 7 | - | 1,418 |
| 1989 | - | - | - | - | - - | - | - | 2,039 |  | - | - | 482 | - | + | - | 2,521 |
| 1990 | - | 7 | - | - | - - | - | - | 1,304 |  | - | - | $321{ }^{2}$ | - | - | - | 1,632 |
| 1991 | 164 | - | - | - | - - | - | - | 2,029 |  | - | - | $522^{2}$ | - | - | - | 2,715 |
| 1992 | - | - | + | - | - - | - | - | 2,349 |  | - | - | 467 | - | - | - | 2,816 |
| 1993 | - | 32 | - | - | - - | 56 | - | 1,754 |  | - | - | 867 | - | - | - | 2,709 |
| 1994 | - | 17 | 217 | - | - - | 15 | - | 1,165 |  | - | - | 175 | - | + | - | 1,589 |
| 1995 | - | 12 | - | - | - - | 25 | - | 1,352 |  | - | - | 270 | 84 | - | - | 1,743 |
| 1996 | - | 2 | + | - | - - | 70 | - | 911 |  | - | - | 198 | - | + | - | 1,181 |
| 1997 | - | 15 | - | - | - - | 62 | - | 610 |  | - | - | 170 | - ${ }^{2}$ | + | - | 857 |
| 1998 | - | 47 | + | - | - - | 23 | - | 859 |  | - | - | 491 | - ${ }^{2}$ | 2 | - | 1,422 |
| 1999 | - | 91 | - | - | - 13 | 7 | - | 1,101 |  | - | - | 1,203 | - ${ }^{2}$ | + | - | 2,415 |
| 2000 | - | - | + | - | - - | 16 | - | 1,021 |  | + | - | 1,169 | $-{ }^{2}$ | 1 | - | 2,206 |
| 2001 | - | - | - | - | - - | 9 | - | $925^{2}$ | + | + | - | 951 | - 2 | 2 | - | 1,887 |
| 2002 | - | - | 3 | - | - - | + | - | $791{ }^{2}$ |  | - | - | 1,167 | $-2$ | + | - | 1,961 |
| 2003 | - | 48 | + | + | + 2 | + | 1 | $949^{2}$ | 1 | 1 | - | 735 | $+^{2}$ | + | + | 1,736 |
| $2004{ }^{1}$ | - | - | - | - | - | + | - | $812^{2}$ |  | - | - | 633 | $-^{2}$ | 3 | - | 1,449 |
| $2005{ }^{1}$ | - | - | - | 1 | - | - | - | $572^{3}$ |  | - | - | 595 | - ${ }^{2}$ | 3 | - | 1,171 |
| $2006{ }^{1}$ | - | 17 | 1 | - | - | 1 | - | $575{ }^{3}$ |  | - | - | 626 | -2 | 2 | - | 1,222 |
| $2007{ }^{1}$ | - | 18 | + | + | + | 3 | - | $514^{3}$ |  | - | - | 438 | + | + | - | 973 |
| $2008{ }^{1}$ | - | 12 | - | 1 | - | 5 | - | 5993 |  | - | - | 390 | - | - | - | 1,007 |
| $2009{ }^{1}$ | - | - | - |  | - - | - | - | $739{ }^{3}$ |  | - | 2 | $483{ }^{2}$ | - | - | - | 1,224 |

[^8]Table 8.3. GREENLAND HALIBUT in Sub areas I and II. Nominal catch ( $t$ ) by countries in Division IIa as officially reported to ICES.

| Year | Esto- <br> nia | Faroe Island s | Fed. <br> Rep. <br> Germ. | France | $\begin{array}{ll} \hline \text { Green } & \text { Ic } \\ - & 1 \\ \text { land } & d \end{array}$ | $\begin{aligned} & \text { Ice- } \\ & \text { lan } \\ & \text { d } \end{aligned}$ | $\begin{aligned} & \text { Ire- } \\ & \text { lan } \\ & \mathrm{d} \end{aligned}$ |  | Norway | Polan d |  | Portu- R al | ussia ${ }^{5}$ | Spain |  | \& W) | UK (Scot.) | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | - | - | 265 | 138 | - |  |  | - | 3,703 |  | - | - | 5,459 | - |  | 1 | - | 9,566 |
| 1985 | - | - | 254 | 239 | - |  |  |  | 4,791 |  | - | - | 6,894 | - |  | 2 | - | 12,180 |
| 1986 | - | 6 | 97 | 13 | - |  | - | - | 6,389 |  | - | - | 5,553 | - |  | 5 | 1 | 12,064 |
| 1987 | - | - | 75 | 13 | - |  | - | - | 5,705 |  | - | - | 4,739 |  |  | 44 | 10 | 10,586 |
| 1988 | - | 177 | 150 | 67 | - |  | - | - | 7,859 |  | - | - | 4,002 | - |  | 56 | 2 | 12,313 |
| 1989 | - | 67 | 104 | 31 | - |  | - | - | 8,050 |  | - | - | 4,964 | - |  | 6 | - | 13,222 |
| 1990 | - | 133 | 12 | 49 | - |  | - | - | 8,233 |  | - | - | 1,246 ${ }^{2}$ | 2 |  | 1 | - | 9,674 |
| 1991 | 1,400 | 314 | 21 | 119 | - |  | - | - | 11,189 |  | - | - | $305^{2}$ |  |  | + | 1 | 13,349 |
| 1992 | - | 16 | 1 | 108 | $13^{4}$ |  | - | - | 3,586 |  | - | $15^{3}$ | 58 | - |  | 1 | - | 3,798 |
| 1993 | - | 29 | 14 | 78 | $8^{4}$ |  | - | - | 7,977 |  | - | 17 | 210 | - |  | 2 | - | 8,335 |
| 1994 | - | - | 33 | 47 | $3^{4}$ |  | 4 | 4 | 6,382 |  | - | 26 | 67 | + |  | 14 | - | 6,576 |
| 1995 | - | - | 30 | 174 | $12^{4}$ |  | 2 | 2 | 6,354 |  | - | 60 | 227 | - |  | 83 | 2 | 6,944 |
| 1996 | - | - | 34 | 219 | 1234 |  | - | - | 9,508 |  | - | 55 | 466 | 4 |  | 278 | 57 | 10,744 |
| 1997 | - | - | 23 | 253 | -4 |  | - | - | 5,702 |  | - | 41 | 334 | $1{ }^{2}$ |  | 21 | 25 | 6,400 |
| 1998 | - | - | 16 | 67 | -4 |  | 1 | 1 | 6,661 |  | - | 80 | 530 | $5^{2}$ |  | 74 | 41 | 7,475 |
| 1999 | - | - | 20 | - | $25^{4}$ |  | 2 | 2 | 13,064 |  | - | 33 | 734 | $1{ }^{2}$ |  | 63 | 45 | 13,987 |
| 2000 | - | - | 10 | 43 | -4 |  | + | + | 7,536 |  | - | 18 | 690 | $1{ }^{2}$ |  | 65 | 43 | 8,406 |
| 2001 | - | - | 49 | 122 | -4 | 9 | - 1 | 1 | 8,740 |  | - | 13 | 726 | $5^{2}$ |  | 56 | 30 | 9,751 |
| 2002 | - | - | 9 | 7 | $22^{4}$ | 4 | 4 | - | 5,780 ${ }^{2}$ |  | - | 3 | 849 | $-2$ |  | 12 | 28 | 6,714 |
| 2003 | - | 390 | 5 | 2 | $12^{4}$ | + | + | + | 6,778 ${ }^{2}$ |  | + | 10 | 1,762 | $14^{2}$ |  | 5 | 58 | 9,036 |
| 2004 | - | - | 4 | - | -4 | 9 | 9 | - | 11,633 ${ }^{2}$ |  | - | 24 | 810 | $4^{2}$ |  | 1 | - | 12,485 |
| 2005 | - | - | 3 | 31 | -4 | - | - | - | 11,216 ${ }^{3}$ |  | - | 11 | 1,406 | + | + | 5 | 18 | 12,690 |
| $2006{ }^{1}$ | - | 175 | - | 38 | - | 7 |  | - | 8,897 ${ }^{3}$ | -2 | 2 | 6 | 950 | + | + | 2 | - | 10,075 |
| $2007{ }^{1}$ | - | 162 | 2 | 37 | + | 12 | , | - | 6,760 ${ }^{3}$ | 2 | 2 | 2 | $489{ }^{2}$ | - |  | + | + | 7,463 |
| $2008{ }^{1}$ | - | 626 | 4 | 38 | - | 23 | , | - | 5,566 ${ }^{3}$ |  | 1 | 1 | 1,170 | 3 |  | 16 | - | 7,448 |
| $2009{ }^{1}$ | - | - | + | - | - | - | - | - | 6,146 ${ }^{3}$ |  | - | 9 | 1,531 ${ }^{2}$ | - |  | 60 | - | 7,746 |

${ }^{1}$ Provisional figures. ${ }^{2}$ Working Group figure. ${ }^{3}$ As reported to Norwegian authorities.
${ }^{4}$ Includes Division Iib. ${ }^{5}$ USSR prior to 1991.

Table 8.4. GREENLAND HALIBUT in Sub-areas I and II. Nominal catch ( $\mathbf{t}$ ) by countries in Division IIb as officially reported to ICES.

| Year | Denmark | Estonia | Faroe Isl. | France | Fed. Re p. Germ. | Ire- <br> land | Lithua -nia | Norway | Po- <br> land | Portugal | Russia ${ }^{4}$ | Spain | UK <br> (E\&W) | $\begin{gathered} \text { UK } \\ \text { (Scot.) } \end{gathered}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | - | - | - | - | 1,900 | - | - - | 80 | - | - | 9,641 | - | 5 | - | 11,626 |
| 1985 | - | - | - | - | 3,746 | - | - - | 71 | - | - | 3,221 | - | 2 | - | 7,040 |
| 1986 | - | - | 36 | - | 2,620 |  | - - | 944 | - | - | 6,032 | - | + | - | 9,632 |
| 1987 | + | - | - | - | 1,947 | - | - - | 572 | - | - | 4,735 | - | 7 | 10 | 7,271 |
| 1988 | - | - | - | - | 590 | - | - - | 239 | - | - | 5,008 | - | 19 | + | 5,856 |
| 1989 | - | - | - | - | 496 | - | - - | 533 | - | - | 3,366 | - | - | - | 4,395 |
| 1990 | - | - | $23^{2}$ | - | 942 | - | - - | 7,706 | - | - | 3,197 ${ }^{2}$ | - | 9 | - | 11,877 |
| 1991 | 11 | 1,000 | - | - | 80 | - | - - | 14,369 | - | - | 1,663 ${ }^{2}$ | 132 | + | 1 | 17,256 |
| 1992 | - | - | - | $3^{2}$ | 12 | - | - - | 1,732 | - | 16 | 193 | 23 | 9 | - | 1,988 |
| 1993 | $2^{3}$ | - | - | $2^{3}$ | 8 | - | $30^{3}$ | 649 | - | 26 | 158 | - | 14 | - | 889 |
| 1994 | 4 | - | $1^{3}$ | $8^{3}$ | 46 | 1 | $4^{3}$ | 881 | - | 10 | 41 | 1 | 62 | 2 | 1,061 |
| 1995 | - | - | - | - | 5 | - | - - | 1,662 | - | 24 | 297 | 1,022 | 32 | 5 | 3,047 |
| 1996 | + | - | - |  | 47 | - | - - | 1,204 | - | 24 | 912 | 196 | 39 | + | 2,422 |
| 1997 | - | - | 12 | - | 33 | 2 | - | 1,349 | 12 | 9 | 534 | $156^{2}$ | 46 | + | 2,153 |
| 1998 | - | - | 10 | - | 18 | 1 | - | 915 | 31 | 19 | 1,638 | $254{ }^{2}$ | 106 | 4 | 2,996 |
| 1999 | - | - | 3 | - | - 14 | - | - - | 839 | 8 | 16 | 1,886 | $318^{2}$ | 31 | - | 3,115 |
| 2000 | - | - | - | 2 | 5 | - | - - | 526 | 3 | 19 | 2,709 | $374{ }^{2}$ | 46 | - | 3,685 |
| 2001 | - | - | - | + | + 9 | - | - - | 1,231 ${ }^{2}$ | 2 | 22 | 3,017 | $413^{2}$ | 42 | - | 4,736 |
| 2002 | - | 219 | - | + | + 30 | 6 | 6 - | $440^{2}$ | 5 | 11 | 3,568 | $178{ }^{2}$ | 29 | - | 4,486 |
| 2003 | + | + | 21 | - | - 13 | - | - - | $620^{2}$ | 4 | 9 | 1,887 | 216 | 35 | + | 2,805 |
| 2004 | - | - | - | - | - 5 | - | - - | 1,395 ${ }^{2}$ | 1 | 26 | 3,219 | $182^{2}$ | 39 | - | 4,866 |
| 2005 | - | 170 | - | - | 5 | - | - - | 1,223 ${ }^{3}$ | - | 12 | 2,882 | $660^{2}$ | 21 | - | 4,973 |
| $2006{ }^{1}$ | - | - | 12 | 8 | 7 | - | 196 | 1,647 ${ }^{3}$ | $201{ }^{2}$ | 20 | 4,479 | $27^{2}$ | 2 | - | 6,600 |
| $2007{ }^{1}$ | - | - | 23 | 3 | 36 | + | + - | $955^{3}$ | $200^{2}$ | 45 | 5,557 | $11^{2}$ | + | + | 6,800 |
| $2008{ }^{1}$ | - | - | 2 - | 3 | 1 | - | - - | 1,229 ${ }^{3}$ | 200 | 45 | 3,734 | 109 | 0 | - | 5,323 |
| $2009{ }^{1}$ | - | - | - | - | $22^{2}$ | - | - - | 1,657 ${ }^{3}$ | 204 | 16 | 1,321 ${ }^{2}$ | 210 | 8 | - | 3,437 |

${ }^{1}$ Provisional figures.
${ }^{2}$ Working Group figure.
${ }^{3}$ As reported to Norwegian authorities.
${ }^{4}$ USSR prior to 1991.

Table 8.5. GREENLAND HALIBUT in the Sub-areas I and II. Landings by gear (tonnes). Approximate figures, the total may differ slightly from Table 8.1

| Year | Gillnet | Longline | Trawl | Danish seine | Onher |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 1189 | 336 | 11759 |  |  |  | 13284 |
| 1981 | 730 | 459 | 13829 |  |  |  | 15018 |
| 1982 | 748 | 679 | 15362 |  |  |  | 16789 |
| 1983 | 1648 | 1388 | 19111 |  |  |  | 22147 |
| 1984 | 1200 | 1453 | 19230 |  |  |  | 21883 |
| 1985 | 1668 | 750 | 17527 |  |  |  | 19945 |
| 1986 | 1677 | 497 | 20701 |  |  |  | 22875 |
| 1987 | 2239 | 588 | 16285 |  |  |  | 19112 |
| 1988 | 2815 | 838 | 15934 |  |  |  | 19587 |
| 1989 | 1342 | 197 | 18599 |  |  |  | 20138 |
| 1990 | 1372 | 1491 | 20325 |  |  |  | 23188 |
| 1991 | 1904 | 4552 | 26864 |  |  |  | 33320 |
| 1992 | 1679 | 1787 | 5787 |  |  |  | 9253 |
| 1993 | 1497 | 2493 | 7889 |  |  |  | 11879 |
| 1994 | 1403 | 2392 | 5353 |  |  |  | 9148 |
| 1995 | 1500 | 4034 | 5494 |  |  |  | 11028 |
| 1996 | 1480 | 4616 | 7977 |  |  |  | 14073 |
| 1997 | 998 | 3378 | 5198 |  |  |  | 9574 |
| 1998 | 1327 | 3891 | 6664 |  |  |  | 11882 |
| 1999 | 2565 | 6804 | 10177 |  |  |  | 19546 |
| 2000 | 1707 | 5029 | 7700 |  |  |  | 14437 |
| 2001 | 2041 | 6303 | 7968 |  |  |  | 16312 |
| 2002 | 1737 | 5309 | 6115 |  |  |  | 13161 |
| 2003 | 2046 | 5483 | 6049 |  |  |  | 13578 |
| 2004 | 2290 | 7135 | 8778 |  | 599 |  | 18801 |
| 2005 | 1842 | 7539 | 9420 |  | 447 |  | 19248 |
| 2006 | 1503 | 6146 | 10042 |  | 205 |  | 17896 |
| 2007 | 997 | 4503 | 9618 |  | 119 |  | 15237 |
| 2008 | 901 | 3575 | 9285 |  | 9 | 8 | 13778 |
| 2009 | 1409 | 4952 | 5994 |  | 34 | 18 | 12407 |

Table 8.6. GREENLAND HALIB UT in Sub-areas I and II. Catch per unit effort and total effort.

| Year | USSR catch/hour tra wling ( t ) |  |  | Norway ${ }^{10}$ catch/hour tra wling ( t ) |  | Average CPUE |  | Totaleffort (in '000 hrs trawling) ${ }^{5}$ | $\begin{gathered} \text { CPUE } \\ 7+{ }^{6} \end{gathered}$ | GDR ${ }^{7}$ <br> (catch/day tonnage (kg) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{RT}^{1}$ |  | PST ${ }^{2}$ | $\mathrm{A}^{8}$ | $\mathrm{B}^{9}$ | $\mathrm{A}^{3}$ | $\mathrm{B}^{4}$ |  |  |  |
| 1965 | 0.80 |  | - | - | - | 0.80 | - | - | - | - |
| 1966 | 0.77 |  | - | - | - | 0.77 | - | - | - | - |
| 1967 | 0.70 |  | - | - | - | 0.70 | - | - | - | - |
| 1968 | 0.65 |  | - | - | - | 0.65 | - | - | - | - |
| 1969 | 0.53 |  | - | - | - | 0.53 | - | - | - | - |
| 1970 | 0.53 |  | - | - | - | 0.53 | - | 169 | 0.50 | - |
| 1971 | 0.46 |  | - | - | - | 0.46 | - | 172 | 0.43 | - |
| 1972 | 0.37 |  | - | - | - | 0.37 | - | 116 | 0.33 | - |
| 1973 | 0.37 |  | - | 0.34 | - | 0.36 | - | 83 | 0.36 | - |
| 1974 | 0.40 |  | . | 0.36 | - | 0.38 | - | 100 | 0.36 | - |
| 1975 | 0.39 |  | 0.51 | 0.38 | - | 0.39 | 0.45 | 99 | 0.37 | - |
| 1976 | 0.40 |  | 0.56 | 0.33 | - | 0.37 | 0.45 | 100 | 0.34 | - |
| 1977 | 0.27 |  | 0.41 | 0.33 | - | 0.30 | 0.37 | 96 | 0.26 | - |
| 1978 | 0.21 |  | 0.32 | 0.21 | - | 0.21 | 0.27 | 123 | 0.17 | - |
| 1979 | 0.23 |  | 0.35 | 0.28 | - | 0.26 | 0.32 | 67 | 0.19 | - |
| 1980 | 0.24 |  | 0.33 | 0.32 | - | 0.28 | 0.33 | 47 | 0.25 | - |
| 1981 | 0.30 |  | 0.36 | 0.36 | - | 0.33 | 0.36 | 42 | 0.28 | - |
| 1982 | 0.26 |  | 0.45 | 0.41 | - | 0.34 | 0.43 | 39 | 0.37 | - |
| 1983 | 0.26 |  | 0.40 | 0.35 | - | 0.31 | 0.38 | 58 | 0.32 | - |
| 1984 | 0.27 |  | 0.41 | 0.32 | - | 0.30 | 0.37 | 59 | 0.30 | - |
| 1985 | 0.28 |  | 0.52 | 0.37 | - | 0.33 | 0.45 | 44 | 0.37 | - |
| 1986 | 0.23 |  | 0.42 | 0.37 | - | 0.30 | 0.40 | 57 | 0.32 | - |
| 1987 | 0.25 |  | 0.50 | 0.35 | - | 0.30 | 0.43 | 44 | 0.35 | - |
| 1988 | 0.20 |  | 0.30 | 0.31 | - | 0.26 | 0.31 | 63 | 0.26 | 4.26 |
| 1989 | 0.20 |  | 0.30 | 0.26 | - | 0.23 | 0.28 | 73 | 0.19 | 2.95 |
| 1990 | - |  | 0.20 | 0.27 | - | - | 0.24 | 95 | 0.16 | 1.66 |
| 1991 | - |  | , | 0.24 | - | - | - | 134 | 0.18 | - |
| 1992 | - |  | - | 0.46 | 0.72 | - | - | 20 | 0.29 | - |
| 1993 | - |  | - | 0.79 | 1.22 | - | - | 15 | 0.65 | - |
| 1994 | - |  | - | 0.77 | 1.27 | - | - | 11 | 0.70 | - |
| 1995 | - |  | - | 1.03 | 1.48 | - | - | - | - | - |
| 1996 | - |  | - | 1.45 | 1.82 | - | - | - | - | - |
| 1997 | 0.71 |  | - | 1.23 | 1.60 | - | - | - | - | - |
| 1998 | 0.71 |  | - | 0.98 | 1.35 | - | - | - | - | - |
| 1999 | 0.84 |  | - | 0.82 | 1.77 | - | - | - | - | - |
| 2000 | 0.94 |  | - | 1.38 | 1.92 | - | - | - | - | - |
| 2001 | 0.82 | ${ }^{11}$ | - | 1.18 | 1.57 | - | - | - | - | - |
| 2002 | 0.85 |  | - | 1.07 | 1.82 | - | - | - | - | - |
| 2003 | 0.97 | ${ }^{12}$ | - | 0.86 | 2.45 | - | - | - | - | - |
| 2004 | 0.63 | 13 | - | 1.16 | 1.79 | - | - | - | - | - |
| 2005 | 0.61 | 12 | - | 1.30 | 2.29 | - | - | - | - | - |
| 2006 | 0.57 | ${ }^{12}$ | - | 0.96 | 2.09 | - | - | - | - | - |
| 2007 | 0.64 | ${ }^{12}$ | - | - | - | - | - | - | - | - |
| 2008 | 0.48 | ${ }^{12}$ | - | - | - | - | - | - | - | - |
| 2009 | 0.77 | 13 | - | - | - | - | - | - | - | - |

${ }^{1}$ Side trawlers, 800-1000 hp. From 1983 onwards, side trawlers (SRTM), 1,000 hp. From 1997 based on research fishing.
2 Stern trawlers, up to $2,000 \mathrm{HP}$.
${ }^{3}$ Arithmetic average of CPUE from USSR RT (or SRT M trawlers) and Norwegian trawlers.
4 Arithmetic average of CPUE from USSR PST and Norwegian trawlers.
${ }^{5}$ For the years 1981-1990, based on average CPUE type B. For 1991-1993, based on the Norwegian CPUE, type A.
${ }^{6}$ Total catch ( $\mathbf{t}$ ) of seven years and older fish divided by total effort.
7 For the years 1988-1989, frost-trawlers 995 BRT (FAO Code 095). For 1990, factory trawlers FVS IV, 1943 BRT (FAO Code 090).
8 Norwegian trawlers, ISSCFV-code 07, 250-499.9 GRT.
9 Norwegian factory trawlers, ISSCFV-code 09, 1000-1999.9 GRT.
${ }^{10}$ From 1992 based on research fishing. 1992-1993: two weeks in May/June and October; 1994-1995: 10 days in May/June.
${ }^{11}$ Based on fishery from april-october only, a period with relatively low CPUE. In previous years fishery was carried out throughout the whole year.
${ }^{12}$ Based on fishery from october-december only, a period with relatively high CPUE
${ }^{13}$ Based on fishery from october-november only.

Table 8.7. Catch numbers at age Numbers* ${ }^{*} 0^{* *}-3$

Run title : NEA Greenland halibut (run: 2010/1)

At 26/04/2010 18:36

Table 1 Catch numbers at age Numbers* $10^{* *}-3$

| YEAR | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| AGE |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 372 | 253 | 170 | 156 | 114 | 1064 | 526 | 80 | 1109 | 212 | 917 |
|  | 1480 | 853 | 563 | 332 | 283 | 2420 | 2792 | 4486 | 3521 | 1117 | 2519 |
|  | 2808 | 1735 | 1106 | 623 | 452 | 3208 | 10464 | 12712 | 9605 | 3923 | 6204 |
|  | 5674 | 3868 | 2715 | 2006 | 1976 | 6288 | 18562 | 12283 | 6438 | 3515 | 3838 |
|  | 4951 | 4203 | 4054 | 3237 | 3923 | 4921 | 10034 | 6130 | 2775 | 2551 | 1834 |
|  | 3981 | 3799 | 2499 | 2409 | 2950 | 4431 | 6671 | 4339 | 1734 | 1919 | 1942 |
|  | 1853 | 1799 | 1284 | 1718 | 2234 | 2381 | 2517 | 2703 | 1368 | 1536 | 1622 |
|  | 1018 | 1002 | 783 | 871 | 792 | 812 | 1250 | 1660 | 1234 | 1127 | 1338 |
|  | 364 | 372 | 246 | 315 | 146 | 229 | 616 | 1044 | 675 | 716 | 734 |
|  | 251 | 282 | 261 | 155 | 43 | 100 | 1104 | 300 | 200 | 251 | 531 |
| +gp | 76 | 50 | 28 | 19 | 7 | 30 | 281 | 143 | 80 | 126 | 216 |
| TOTALNUM | 22828 | 18216 | 13709 | 11841 | 12920 | 25884 | 54817 | 45880 | 28739 | 16993 | 21695 |
| TONSLAND | 40391 | 34751 | 26321 | 24267 | 26168 | 43789 | 89484 | 79034 | 43055 | 29938 | 37763 |
| SOPCOF \% | 100 | 100 | 101 | 100 | 100 | 103 | 94 | 104 | 98 | 92 | 98 |

Table 1 Catch numbers at age Numbers*10**-3

| YEAR | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| AGE |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 840 | 830 | 2037 | 1897 | 2218 | 731 | 1896 | 1304 | 1543 | 915 | 1219 |
| 6 | 2337 | 2982 | 3255 | 3589 | 3155 | 1138 | 1917 | 1494 | 1864 | 3698 | 2874 |
| 7 | 6520 | 5824 | 4200 | 4118 | 2727 | 1665 | 1919 | 1276 | 1851 | 3350 | 2561 |
| 8 | 4118 | 5002 | 2524 | 2365 | 1234 | 1341 | 933 | 1208 | 2287 | 1938 | 1548 |
| 9 | 2265 | 3000 | 1610 | 1509 | 495 | 944 | 484 | 1493 | 1491 | 1064 | 972 |
| 10 | 1654 | 1350 | 1104 | 946 | 319 | 473 | 448 | 1258 | 1228 | 1191 | 1037 |
| 11 | 1857 | 915 | 1062 | 934 | 296 | 511 | 482 | 838 | 713 | 602 | 614 |
| 12 | 1536 | 1212 | 858 | 438 | 243 | 275 | 380 | 502 | 488 | 340 | 363 |
| 13 | 1122 | 698 | 595 | 349 | 103 | 242 | 384 | 324 | 247 | 171 | 161 |
| 14 | 600 | 526 | 384 | 147 | 45 | 145 | 150 | 108 | 201 | 132 | 120 |
| +gp | 368 | 358 | 180 | 112 | 51 | 78 | 62 | 46 | 64 | 71 | 63 |
| 0 TOTALNUM | 23217 | 22697 | 17809 | 16404 | 10886 | 7543 | 9055 | 9851 | 11977 | 13472 | 11532 |
| TONSLAND | 38172 | 36074 | 28827 | 24617 | 17312 | 13284 | 15018 | 16789 | 22147 | 21883 | 19945 |
| SOPCOF \% | 88 | 93 | 101 | 105 | 104 | 109 | 107 | 100 | 98 | 100 | 99 |

Table 8.7 (Continued)

Table 1 Catch numbers at age Numbers*10**-3

| YEAR | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

AGE

| 5 | 1672 | 1212 | 907 | 2080 | 2139 | 3312 | 1098 | 1140 | 631 | 846 | 1034 | 330 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | 3335 | 2972 | 2540 | 4453 | 5163 | 3889 | 1195 | 1088 | 708 | 992 | 2083 | 921 |
| 7 | 2712 | 3572 | 3141 | 3655 | 4642 | 4716 | 1069 | 1608 | 1252 | 1719 | 3795 | 1822 |
| 8 | 1531 | 1746 | 2096 | 1657 | 1932 | 2355 | 778 | 1118 | 817 | 990 | 1426 | 953 |
| 9 | 1128 | 752 | 1182 | 801 | 1221 | 1031 | 360 | 140 | 310 | 405 | 262 | 342 |
| 10 | 997 | 828 | 860 | 318 | 499 | 1284 | 600 | 976 | 642 | 726 | 655 | 822 |
| 11 | 530 | 362 | 481 | 228 | 264 | 774 | 188 | 444 | 416 | 461 | 270 | 231 |
| 12 | 434 | 202 | 313 | 126 | 314 | 673 | 150 | 144 | 330 | 371 | 132 | 150 |
| 13 | 314 | 186 | 133 | 120 | 42 | 177 | 79 | 36 | 88 | 154 | 29 | 18 |
| 14 | 305 | 63 | 140 | 140 | 96 | 266 | 89 | 20 | 39 | 56 | 22 | 41 |
| +gp | 239 | 7 | 47 | 28 | 44 | 517 | 56 | 4 | 3 | 8 | 1 | 1 |
| 0 TOTALNUM | 13197 | 11902 | 11840 | 13606 | 16356 | 18994 | 5662 | 6718 | 5236 | 6728 | 9709 | 5631 |
| TONSLAND | 22875 | 19112 | 19587 | 20138 | 23183 | 33320 | 8602 | 11933 | 9226 | 11734 | 14347 | 9410 |
| SOPCOF \% | 98 | 101 | 100 | 103 | 102 | 105 | 95 | 102 | 99 | 101 | 101 | 99 |

Table 1 Catch numbers at age Numbers*10**-3
$\begin{array}{lllllllllllllllllllllllll}\text { YEAR } & 1998 & 1999 & 2000 & 2001 & 2002 & 2003 & 2004 & 2005 & 2006 & 2007 & 2008 & 2009\end{array}$

| AGE |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 359 | 433 | 380 | 441 | 277 | 397 | 290 | 429 | 548 | 987 | 449 | 959 |
| 6 | 1116 | 1905 | 735 | 1347 | 921 | 1025 | 1016 | 1072 | 1347 | 1598 | 751 | 1137 |
| 7 | 2466 | 3955 | 1926 | 2338 | 1475 | 1827 | 2316 | 1962 | 2067 | 2202 | 1231 | 1384 |
| 8 | 1464 | 1810 | 1464 | 1325 | 983 | 928 | 1392 | 1766 | 1584 | 1134 | 1277 | 1746 |
| 9 | 527 | 914 | 743 | 788 | 631 | 632 | 1087 | 936 | 1034 | 629 | 790 | 723 |
| 10 | 924 | 1905 | 1318 | 1140 | 1097 | 1045 | 778 | 991 | 691 | 436 | 314 | 255 |
| 11 | 237 | 380 | 457 | 519 | 563 | 520 | 675 | 616 | 485 | 426 | 365 | 514 |
| 12 | 122 | 237 | 330 | 372 | 301 | 311 | 607 | 622 | 548 | 464 | 412 | 325 |
| 13 | 15 | 67 | 49 | 115 | 132 | 77 | 199 | 376 | 466 | 246 | 341 | 300 |
| 14 | 29 | 42 | 37 | 54 | 59 | 107 | 155 | 244 | 209 | 169 | 207 | 96 |
| +gp | 15 | 7 | 14 | 12 | 42 | 26 | 105 | 328 | 230 | 224 | 247 | 115 |
| 0 TOTALNUM | 7274 | 11655 | 7453 | 8451 | 6481 | 6895 | 8620 | 9342 | 9209 | 8515 | 6384 | 7554 |
| TONSLAND | 11893 | 19517 | 14437 | 16307 | 13161 | 13578 | 18800 | 18834 | 17897 | 15237 | 13778 | 12407 |
| SOPCOF \% | 100 | 102 | 101 | 100 | 100 | 100 | 99 | 97 | 100 | 96 | 101 | 102 |

Table 8.8. Catch weights at age (kg)
Run title : NEA Greenland halibut (run: 2010/1)

At 26/04/2010 1836

| YEAR | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
| 5 | 0.42 | 0.42 | 0.42 | 0.42 | 0.42 | 0.42 | 0.567 | 0.567 | 0.567 | 0.567 | 0.567 |
| 6 | 0.64 | 0.64 | 0.64 | 0.65 | 0.66 | 0.64 | 0.737 | 0.737 | 0.737 | 0.737 | 0.737 |
| 7 | 0.9 | 0.9 | 0.91 | 0.93 | 0.96 | 0.91 | 1.079 | 1.079 | 1.079 | 1.079 | 1.079 |
| 8 | 1.2 | 1.22 | 1.24 | 1.27 | 1.31 | 1.25 | 1.421 | 1.421 | 1.421 | 1.421 | 1.421 |
| 9 | 1.63 | 1.66 | 1.7 | 1.71 | 1.74 | 1.64 | 1.848 | 1.848 | 1.848 | 1.848 | 1.848 |
| 10 | 2.26 | 2.23 | 2.22 | 2.2 | 2.19 | 2.25 | 2.281 | 2.281 | 2.281 | 2.281 | 2.281 |
| 11 | 3.11 | 3 | 2.94 | 2.84 | 2.79 | 2.99 | 2.887 | 2.887 | 2.887 | 2.887 | 2.887 |
| 12 | 3.74 | 3.49 | 3.39 | 3.3 | 3.19 | 3.63 | 3.247 | 3.247 | 3.247 | 3.247 | 3.247 |
| 13 | 4.57 | 4.4 | 4.38 | 4.27 | 4.27 | 4.68 | 4.303 | 4.303 | 4.303 | 4.303 | 4.303 |
| 14 | 5.01 | 4.91 | 4.84 | 4.88 | 5 | 5.38 | 4.931 | 4.931 | 4.931 | 4.931 | 4.931 |
| +gp | 5.94 | 5.89 | 5.88 | 5.8 | 5.99 | 5.99 | 5.794 | 5.841 | 6.037 | 6.006 | 5.964 |
| 0 SOPCOF AC | 0.9986 | 1.0046 | 1.0054 | 1.0024 | 0.9994 | 1.0262 | 0.9436 | 1.0434 | 0.9752 | 0.9231 | 0.9825 |

Table 2 Catch weights at age ( kg )

| YEAR | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 5 | 0.567 | 0.567 | 0.567 | 0.567 | 0.9 | 0.702 | 0.66 | 0.69 | 0.75 | 0.63 | 0.6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | 0.737 | 0.737 | 0.737 | 0.737 | 1.2 | 0.872 | 0.84 | 0.84 | 1.04 | 0.96 | 0.89 |
| 7 | 1.079 | 1.079 | 1.079 | 1.079 | 1.5 | 1.141 | 1.15 | 1.03 | 1.34 | 1.18 | 1.2 |
| 8 | 1.421 | 1.421 | 1.421 | 1.421 | 1.8 | 1.468 | 1.56 | 1.31 | 1.57 | 1.53 | 1.85 |
| 9 | 1.848 | 1.848 | 1.848 | 1.848 | 2.2 | 1.778 | 2.04 | 1.74 | 1.97 | 2.31 | 2.59 |
| 10 | 2.281 | 2.281 | 2.281 | 2.281 | 2.6 | 2.302 | 2.57 | 2.24 | 2.73 | 2.87 | 3.18 |
| 11 | 2.887 | 2.887 | 2.887 | 2.887 | 3 | 2.664 | 2.98 | 2.77 | 3.29 | 3.46 | 3.62 |
| 12 | 3.247 | 3.247 | 3.247 | 3.247 | 3.5 | 3.046 | 3.43 | 3.37 | 4.22 | 3.77 | 3.95 |
| 13 | 4.303 | 4.303 | 4.303 | 4.303 | 4.1 | 3.368 | 4.13 | 4.32 | 4.71 | 3.99 | 4.48 |
| 14 | 4.931 | 4.931 | 4.931 | 4.931 | 4.8 | 4.285 | 4.68 | 5.35 | 6.08 | 4.35 | 4.25 |
| +gp | 5.91 | 5.923 | 6.027 | 5.906 | 6.176 | 5.346 | 5.999 | 5.833 | 6.122 | 4.525 | 4.825 |
| 0 SOPCOF AC | 0.8805 | 0.9255 | 1.0095 | 1.0485 | 1.0364 | 1.0894 | 1.068 | 1.0038 | 0.9783 | 1.0009 | 0.9858 |

## Table 8.8 (Continued)

| YEAR | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
| 5 | 0.62 | 0.709 | 0.74 | 0.76 | 0.71 | 0.77 | 0.68 | 0.79 | 0.72 | 0.73 | 0.77 |
| 6 | 0.92 | 1.003 | 0.962 | 1.03 | 1.06 | 1.05 | 0.97 | 1.02 | 0.94 | 0.94 | 0.97 |
| 7 | 1.28 | 1.266 | 1.249 | 1.32 | 1.29 | 1.38 | 1.27 | 1.35 | 1.27 | 1.25 | 1.31 |
| 8 | 1.9 | 1.683 | 1.626 | 1.8 | 1.7 | 1.75 | 1.76 | 1.88 | 1.72 | 1.74 | 1.74 |
| 9 | 2.48 | 2.482 | 2.164 | 2.42 | 2.1 | 2.2 | 2.21 | 2.46 | 2.19 | 2.09 | 2.24 |
| 10 | 3.11 | 2.982 | 2.897 | 3.13 | 2.61 | 2.6 | 2.56 | 2.67 | 2.52 | 2.51 | 2.59 |
| 11 | 3.35 | 3.547 | 3.406 | 3.37 | 2.87 | 2.79 | 3.11 | 3.43 | 2.97 | 2.95 | 3.29 |
| 12 | 3.72 | 3.8 | 3.661 | 4.05 | 3.45 | 3.28 | 3.59 | 4.29 | 3.29 | 3.34 | 4.02 |
| 13 | 4 | 4.56 | 4.247 | 4.29 | 3.72 | 3.89 | 3.83 | 5.08 | 3.84 | 3.83 | 4.75 |
| 14 | 4.18 | 5.002 | 4.187 | 4.5 | 4.09 | 4.38 | 4.25 | 6.33 | 4.95 | 4.98 | 6.24 |
| +gp | 4.526 | 5.953 | 4.463 | 4.72 | 4.52 | 5.29 | 4.8 | 8.91 | 6.68 | 8.15 | 6.09 |
| 0 SOPCOF AC | 0.9782 | 1.0116 | 0.9973 | 1.0346 | 1.0204 | 1.047 | 0.9519 | 1.0183 | 0.9937 | 1.0095 | 1.0066 |


| YEAR | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AGE |  |  |  |  |  |  |  |  |  |  |  |
| 5 | 0.77 | 0.73 | 0.7 | 0.76 | 0.74 | 0.69 | 0.715 | 0.77 | 0.669 | 0.637 | 0.626 |
| 6 | 0.94 | 0.93 | 0.95 | 0.97 | 1.03 | 0.94 | 1.05 | 1.095 | 0.952 | 0.86 | 0.903 |
| 7 | 1.28 | 1.3 | 1.27 | 1.33 | 1.39 | 1.36 | 1.428 | 1.498 | 1.306 | 1.149 | 1.313 |
| 8 | 1.64 | 1.61 | 1.55 | 1.63 | 1.75 | 1.68 | 1.748 | 1.903 | 1.653 | 1.53 | 1.686 |
| 9 | 2.07 | 2.12 | 2 | 2.11 | 2.29 | 2.18 | 2.318 | 2.463 | 2.131 | 2.122 | 2.321 |
| 10 | 2.59 | 2.57 | 2.46 | 2.61 | 2.68 | 2.68 | 2.615 | 2.775 | 2.544 | 2.622 | 2.553 |
| 11 | 3.3 | 3.25 | 3.22 | 3.35 | 3.33 | 3.19 | 3.043 | 3.128 | 2.848 | 2.699 | 2.925 |
| 12 | 4.01 | 3.91 | 3.85 | 3.97 | 3.92 | 3.89 | 3.694 | 3.809 | 3.334 | 3.315 | 3.189 |
| 13 | 4.83 | 4.9 | 4.61 | 4.97 | 4.81 | 4.46 | 4.566 | 4.291 | 3.734 | 3.998 | 3.747 |
| 14 | 5.95 | 5.66 | 5.84 | 5.82 | 5.81 | 5.25 | 5.568 | 5.453 | 4.384 | 4.641 | 4.539 |
| +gp | 6.26 | 4.91 | 5.98 | 7.22 | 7.41 | 6.32 | 6.365 | 6.355 | 5.791 | 6.743 | 9.078 |
| 0 SOPCOF AC | 0.9851 | 0.9983 | 1.0172 | 1.0055 | 1.0014 | 1 | 0.996 | 0.9853 | 0.9655 | 1.0042 | 0.9592 |


| Table 2 Catch weights at age (kg) |  |  |
| :---: | :---: | :---: |
| YEAR | 2008 | 2009 |
| AGE |  |  |
| 5 | 0.695 | 0.567 |
| 6 | 0.919 | 0.802 |
| 7 | 1.359 | 1.071 |
| 8 | 1.756 | 1.471 |
| 9 | 2.231 | 1.928 |
| 10 | 2.378 | 2.216 |
| 11 | 2.855 | 2.63 |
| 12 | 3.23 | 3.082 |
| 13 | 3.546 | 3.791 |
| 14 | 3.915 | 4.528 |
| +gp | 7.453 | 7.069 |
| 0 SOPCOF AC | 1.0086 | 1.0165 |

Table 8.9. Proportion mature at age

Run title : NEA Greenland halibut (run: 2010/1)

At 26/04/2010 1836

Table 5 Proportion mature at age
$\begin{array}{lllllllllllll}\text { YEAR } & 1964 & 1965 & 1966 & 1967 & 1968 & 1969 & 1970 & 1971 & 1972 & 1973 & 1974\end{array}$
AGE

| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 6 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 |
| 7 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 |
| 8 | 0.21 | 0.21 | 0.21 | 0.21 | 0.21 | 0.21 | 0.21 | 0.21 | 0.21 | 0.21 | 0.21 |
| 9 | 0.67 | 0.67 | 0.67 | 0.67 | 0.67 | 0.67 | 0.67 | 0.67 | 0.67 | 0.67 | 0.67 |
| 10 | 0.86 | 0.86 | 0.86 | 0.86 | 0.86 | 0.86 | 0.86 | 0.86 | 0.86 | 0.86 | 0.86 |
| 11 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 |
| 12 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 |
| 13 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 14 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
|  | $+g p$ | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

Table 5 Proportion mature at age $\begin{array}{llllllllllllll}\text { YEAR } & 1975 & 1976 & 1977 & 1978 & 1979 & 1980 & 1981 & 1982 & 1983 & 1984 & 1985\end{array}$ AGE

| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 6 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.04 |
| 7 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 |
| 8 | 0.21 | 0.21 | 0.21 | 0.21 | 0.21 | 0.21 | 0.21 | 0.21 | 0.18 | 0.18 |
| 9 | 0.67 | 0.67 | 0.67 | 0.67 | 0.67 | 0.67 | 0.67 | 0.67 | 0.6 | 0.61 |
| 10 | 0.86 | 0.86 | 0.86 | 0.86 | 0.86 | 0.86 | 0.86 | 0.86 | 0.82 | 0.83 |
| 11 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.96 | 0.97 |
| 12 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 |
| 13 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 14 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

## Table 8.9 (Continued)

Table 5 Proportion mature at age

| YEAR | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | AGE


| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0.03 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0 |
| 0.03 | 0.02 | 0.01 | 0.02 | 0.02 | 0.04 | 0.06 | 0.08 | 0.07 | 0.08 | 0.07 |
| 0.24 | 0.22 | 0.21 | 0.18 | 0.17 | 0.15 | 0.28 | 0.32 | 0.34 | 0.29 | 0.25 |
| 0.74 | 0.66 | 0.53 | 0.49 | 0.51 | 0.54 | 0.66 | 0.68 | 0.69 | 0.58 | 0.58 |
| 0.91 | 0.9 | 0.87 | 0.8 | 0.77 | 0.77 | 0.86 | 0.83 | 0.81 | 0.79 | 0.88 |
| 0.99 | 0.95 | 0.89 | 0.89 | 0.91 | 0.89 | 0.87 | 0.88 | 0.95 | 0.96 | 0.97 |
| 0.98 | 0.98 | 0.98 | 1 | 1 | 1 | 1 | 0.94 | 0.94 | 0.89 | 0.94 |
| 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

Table 5 Proportion mature at age

| YEAR | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | AGE

Table 5 Proportion mature at age
YEAR 20082009
AGE

| 0 | 0 |
| :--- | :--- |
| 0.01 | 0.02 |
| 0.03 | 0.03 |
| 0.07 | 0.12 |
| 0.24 | 0.42 |
| 0.36 | 0.59 |
| 0.58 | 0.79 |
| 0.73 | 0.86 |
| 0.82 | 0.95 |
| 0.96 | 1 |
| 0.99 | 1 |

## Table 8.10. Extended Survivors Analysis

## Lowestoft VPA Version 3.1

## 26/04/2010 18:34

Extended Survivors Analysis
Arctic Green.halibut (run: 2010/1)
CPUE data from file fleet

Catch data for 46 years. 1964 to 2009. Ages 5 to 15 .

| Fleet | First year | Last year | First age |  | Last age |  | Alpha | Beta |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FLT04: Norw. Exp. CP | 1992 | 2009 |  | 5 |  | 14 | 0.38 | 0.44 |
| FLT07: Russ.Surv. ne | 1992 | 2009 |  | 5 |  | 14 | 0.75 | 0.92 |
| FLT08: Norw.Comb.Sur | 1996 | 2009 |  | 5 |  | 14 | 0.55 | 0.72 |
| Time series weights : |  |  |  |  |  |  |  |  |

Tapered time weighting applied
Power = 3 over 20 years

Catchability analysis :
Catchability independent of stock size for all ages
Catchability independent of age for ages $>=10$

Terminal population estimation :
Terminal year survivor estimates shrunk to wards the mean $F$ of the final 2 years.
S.E. of the mean to which the estimates are shrunk $=.500$

Oldest age survivor estimates for the years 1964 to 2009
shrunk towards $1.000^{*}$ the mean F of ages 9-13
S.E. of the mean to which the estimates are shrunk $=.500$

Minimum standard error for population estimates from each cohort age $=.300$

Individual fleet weighting not applied
Tuning converged after 56 iterations

| Regression weights |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 0.751 | 0.82 | 0.877 | 0.921 | 0.954 | 0.976 | 0.99 | 0.997 | 1 | 1 |
| Fishing mortalities |  |  |  |  |  |  |  |  |  |  |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 5 | 0.022 | 0.026 | 0.015 | 0.023 | 0.016 | 0.017 | 0.02 | 0.041 | 0.015 | 0.02 |
| 6 | 0.054 | 0.098 | 0.067 | 0.067 | 0.07 | 0.071 | 0.065 | 0.072 | 0.038 | 0.047 |
| 7 | 0.174 | 0.232 | 0.14 | 0.174 | 0.2 | 0.178 | 0.18 | 0.137 | 0.069 | 0.086 |
| 8 | 0.195 | 0.165 | 0.136 | 0.116 | 0.184 | 0.219 | 0.202 | 0.134 | 0.104 | 0.125 |
| 9 | 0.154 | 0.145 | 0.104 | 0.115 | 0.184 | 0.172 | 0.182 | 0.109 | 0.124 | 0.075 |
| 10 | 0.48 | 0.353 | 0.29 | 0.237 | 0.192 | 0.24 | 0.175 | 0.103 | 0.069 | 0.051 |
| 11 | 0.346 | 0.332 | 0.278 | 0.205 | 0.224 | 0.217 | 0.168 | 0.148 | 0.111 | 0.146 |
| 12 | 0.58 | 0.496 | 0.308 | 0.231 | 0.369 | 0.314 | 0.288 | 0.227 | 0.197 | 0.13 |
| 13 | 0.235 | 0.383 | 0.307 | 0.113 | 0.214 | 0.388 | 0.387 | 0.191 | 0.245 | 0.204 |
| 14 | 0.417 | 0.415 | 0.326 | 0.414 | 0.329 | 0.417 | 0.365 | 0.222 | 0.231 | 0.095 |

## Table 8.10 (Continued)

XSA population numbers (Thousands)

|  |  | AG |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| 2000 | 1.85E+04 | 1. $50 \mathrm{E}+04$ | 1. $30 \mathrm{E}+04$ | 8.90E+03 | $5.60 \mathrm{E}+03$ | $3.72 \mathrm{E}+03$ | 1.68E+03 | $8.08 \mathrm{E}+02$ | 2. $52 \mathrm{E}+02$ | 1.17E+02 |
| 2001 | 1. $83 \mathrm{E}+04$ | 1. $56 \mathrm{E}+04$ | 1. $22 \mathrm{E}+04$ | $9.41 \mathrm{E}+03$ | $6.30 \mathrm{E}+03$ | $4.13 \mathrm{E}+03$ | 1.98E+03 | 1.03E+03 | $3.90 \mathrm{E}+02$ | 1.71E+02 |
| 2002 | 2. $02 \mathrm{E}+04$ | 1. $53 \mathrm{E}+04$ | 1. $22 \mathrm{E}+04$ | $8.33 \mathrm{E}+03$ | $6.87 \mathrm{E}+03$ | 4.70E+03 | $2.50 \mathrm{E}+03$ | 1.22E+03 | $5.38 \mathrm{E}+02$ | 2. $29 \mathrm{E}+02$ |
| 2003 | 1.92E+04 | 1. $71 \mathrm{E}+04$ | 1. $23 \mathrm{E}+04$ | 9.12E+03 | $6.25 \mathrm{E}+03$ | $5.33 \mathrm{E}+03$ | 3.02E+03 | 1. $63 \mathrm{E}+03$ | $7.75 \mathrm{E}+02$ | 3.40E+02 |
| 2004 | 1.99E+04 | 1. $62 \mathrm{E}+04$ | 1. $38 \mathrm{E}+04$ | 8. $92 \mathrm{E}+03$ | $6.98 \mathrm{E}+03$ | $4.80 \mathrm{E}+03$ | $3.62 \mathrm{E}+03$ | 2.12E+03 | 1.11E+03 | 5. $96 \mathrm{E}+02$ |
| 2005 | 2.72E+04 | 1. $69 \mathrm{E}+04$ | 1. $30 \mathrm{E}+04$ | $9.69 \mathrm{E}+03$ | $6.39 \mathrm{E}+03$ | $5.00 \mathrm{E}+03$ | 3.41E+03 | $2.49 \mathrm{E}+03$ | 1.26E+03 | 7. $72 \mathrm{E}+02$ |
| 2006 | $2.94 \mathrm{E}+04$ | 2.30E+04 | 1. $35 \mathrm{E}+04$ | $9.35 \mathrm{E}+03$ | $6.71 \mathrm{E}+03$ | $4.63 \mathrm{E}+03$ | 3. $39 \mathrm{E}+03$ | $2.36 \mathrm{E}+03$ | 1. $56 \mathrm{E}+03$ | 7. $37 \mathrm{E}+02$ |
| 2007 | $2.63 \mathrm{E}+04$ | $2.48 \mathrm{E}+04$ | 1. $86 \mathrm{E}+04$ | $9.73 \mathrm{E}+03$ | $6.58 \mathrm{E}+03$ | $4.81 \mathrm{E}+03$ | 3. $35 \mathrm{E}+03$ | $2.47 \mathrm{E}+03$ | 1. $52 \mathrm{E}+03$ | $9.15 \mathrm{E}+02$ |
| 2008 | $3.17 \mathrm{E}+04$ | $2.17 \mathrm{E}+04$ | 1. $99 \mathrm{E}+04$ | 1. $40 \mathrm{E}+04$ | $7.32 \mathrm{E}+03$ | $5.08 \mathrm{E}+03$ | 3.74E+03 | $2.48 \mathrm{E}+03$ | 1. $69 \mathrm{E}+03$ | 1.08E+03 |
| 2009 | $5.27 \mathrm{E}+04$ | $2.69 \mathrm{E}+04$ | 1. $80 \mathrm{E}+04$ | 1. $60 \mathrm{E}+04$ | $1.08 \mathrm{E}+04$ | $5.57 \mathrm{E}+03$ | $4.08 \mathrm{E}+03$ | $2.88 \mathrm{E}+03$ | 1.76E+03 | 1.14E+03 |
| Estimated population abundance at 1st Jan 2010 |  |  |  |  |  |  |  |  |  |  |

Taper weighted geometric mean of the VPA populations:

Standard error of the weighted Log(VPA populations) :
0.3242
0.2405
0.2551
0.3276
0.3757
0.3779
0.5869
0.7462
0.9246
0.9718

Log catchability residuals.
Fleet : FLTO 4: Norw. Exp. CP

| Age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 0.45 | 1.02 | 0.76 | 0.88 | 1.11 | 1.02 | -0.55 | -0.21 |  |  |
| 6 | -0.1 | 0.16 | 0.28 | 0 | 0.82 | 0.22 | -0.12 | -0.1 |  |  |
| 7 | -0.39 | 0.19 | 0.21 | 0.21 | 0.43 | 0.11 | 0.09 | -0.11 |  |  |
| 8 | -0.07 | 0.3 | 0.38 | 0.39 | 0.28 | -0.11 | -0.03 | -0.13 |  |  |
| 9 | -1.49 | -1.47 | -0.97 | 0.23 | -0.27 | -0.07 | -0.26 | -1. 22 |  |  |
| 10 | -0.28 | 0.24 | 0.44 | 0.9 | 0.14 | 0.61 | -0.93 | 0.32 |  |  |
| 11 | -0.06 | 0.02 | -0.07 | 0.32 | -0.54 | 0.62 | -0.88 | -1.03 |  |  |
| 12 | 0.22 | -0.05 | -0.69 | 0.27 | -0.67 | 0.55 | -0.82 | 0.61 |  |  |
| 13 | -0.26 | 0.03 | -0.64 | -0.09 | 99.99 | 0.15 | 99.99 | -0.64 |  |  |
| 14 | -1. 4 | -0.2 | -0.51 | 0.18 | -0.17 | -0.12 | 99.99 | -0.08 |  |  |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 5 | 0.35 | -0.37 | -0.22 | -0.07 | -0.04 | -0.67 | 99.99 | 99.99 | 99.99 | 99.99 |
| 6 | -0.03 | -0.08 | -0.21 | -0.06 | -0.12 | 0.1 | 99.99 | 99.99 | 99.99 | 99.99 |
| 7 | 0.29 | -0.23 | 0.18 | -0.14 | -0.2 | -0.17 | 99.99 | 99.99 | 99.99 | 99.99 |
| 8 | -0.11 | 0.31 | -0.21 | -0.56 | -0.06 | 0.43 | 99.99 | 99.99 | 99.99 | 99.99 |
| 9 | 0 | 0.23 | 0.09 | 0.32 | 0.41 | 0.54 | 99.99 | 99.99 | 99.99 | 99.99 |
| 10 | 0.45 | -0.06 | 0.01 | 0.1 | -0.64 | -0.16 | 99.99 | 99.99 | 99.99 | 99.99 |
| 11 | -1.06 | -0.74 | -0.75 | -0.34 | -0.51 | -0.53 | 99.99 | 99.99 | 99.99 | 99.99 |
| 12 | -0.06 | -0.07 | -0.67 | -0.01 | -0.04 | 0.09 | 99.99 | 99.99 | 99.99 | 99.99 |
| 13 | 0.34 | -0.88 | -1.64 | -0.29 | -0.33 | 0.14 | 99.99 | 99.99 | 99.99 | 99.99 |
| 14 | 99.99 | -0.45 | -0.09 | -0.22 | -0.12 | -0.09 | 99.99 | 99.99 | 99.99 | 99.99 |

Table 8.10 (Continued)
Mean log catchability and standarde rror of ages with catchability independent of year class st rength and constant w.r.t. ti me

| Age | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Mean $\log q$ | -5.2259 | -4.1666 | -3.3703 | -3.8336 | -4.5028 | -3.7661 | -3.7661 | -3.7661 | -3.7661 | -3.7661 |
| S.E $(\log q)$ | 0.5741 | 0.2333 | 0.2211 | 0.3219 | 0.5629 | 0.4777 | 0.7437 | 0.458 | 0.7743 | 0.2778 |

Regression statistics :
Ages with q independent of year class strength and constant w.r.t. time.

| Age | Slope | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 5 | -1.63 | -1.169 | 17.5 | 0.03 | 14 | 0.91 | -5.23 |
| 6 | 1.67 | -0.628 | 0.49 | 0.13 | 14 | 0.41 | -4.17 |
| 7 | 1.42 | -0.696 | 0.86 | 0.32 | 14 | 0.33 | -3.37 |
| 8 | 1.44 | -0.68 | 1.55 | 0.28 | 14 | 0.48 | -3.83 |
| 9 | 0.59 | 1.314 | 6.17 | 0.63 | 14 | 0.32 | -4.5 |
| 10 | 1.6 | -0.876 | 1.11 | 0.26 | 14 | 0.78 | -3.77 |
| 11 | 1.24 | -0.646 | 3.55 | 0.55 | 14 | 0.59 | -4.31 |
| 12 | 0.96 | 0.152 | 3.97 | 0.74 | 14 | 0.46 | -3.86 |
| 13 | 1 | -0.009 | 4.16 | 0.59 | 12 | 0.7 | -4.17 |
| 14 | 0.97 | 0.286 | 3.99 | 0.95 | 12 | 0.21 | -3.94 |
| 1 |  |  |  |  |  |  |  |

Fleet: FLT07: Russ.Surv. ne

| Age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 1.92 | 0.77 | 0.07 | -0.44 | -0.33 | -0.98 | -0.27 | -0.4 |  |  |
| 6 | 1.06 | 0.75 | 0.34 | -0.04 | 0.09 | -0.45 | -0.36 | -0.49 |  |  |
| 7 | 0.61 | 0.63 | 0.13 | 0.11 | 0.17 | -0.2 | -0.24 | -0.47 |  |  |
| 8 | 0.48 | 0.47 | 0.2 | 0.45 | 0.31 | 0.1 | 0.15 | -0.01 |  |  |
| 9 | -0.51 | 0.04 | 0.12 | 0.42 | 0.85 | -0.05 | 0.25 | 0.11 |  |  |
| 10 | -0.27 | 0.15 | 0.42 | 0.36 | -0.71 | 0.13 | 0.31 | 0.22 |  |  |
| 11 | 0.53 | 0.02 | -0.32 | 0.07 | -0.52 | 0.42 | 0.86 | -0.12 |  |  |
| 12 | 0.39 | 0.54 | 0.11 | 0.16 | -0.79 | -0.31 | 0.63 | 0.33 |  |  |
| 13 | -0.35 | -0.23 | -0.28 | -0.19 | -0.35 | 0.49 | 0.48 | 0.65 |  |  |
| 14 | -5.09 | 0.75 | 0.54 | -1.69 | -0.32 | -0.35 | -0.26 | -0.19 |  |  |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 5 | 0.15 | 0.65 | 99.99 | 99.99 | -0.05 | -0.16 | 0.25 | 0.21 | -0.24 | 0.29 |
| 6 | -0.17 | 0.7 | 99.99 | 99.99 | 0.09 | -0.14 | 0.2 | -0.09 | -0.26 | 0.26 |
| 7 | -0.23 | 0.28 | 99.99 | 99.99 | -0.06 | -0.13 | 0.43 | 0.05 | -0.18 | 0.15 |
| 8 | 0.16 | -0.33 | 99.99 | 99.99 | -0.25 | -0.31 | 0.08 | -0.09 | 0.1 | 0.11 |
| 9 | 0.16 | -0.3 | 99.99 | 99.99 | -0.13 | -0.64 | -0.18 | -0.02 | 0.62 | -0.18 |
| 10 | 0.27 | 0.17 | 99.99 | 99.99 | -0.17 | -0.32 | -0.2 | 0.22 | 0.18 | -0.38 |
| 11 | 0.62 | 0.13 | 99.99 | 99.99 | -0.26 | -0.52 | -0.26 | 0.48 | 0.67 | 0.62 |
| 12 | 0.62 | 0.84 | 99.99 | 99.99 | 0.02 | -0.28 | 0.14 | 0.84 | 1.11 | 0.47 |
| 13 | -0.74 | 1.1 | 99.99 | 99.99 | 0.01 | -0.24 | 0.29 | 0.48 | 1.29 | 0.94 |
| 14 | 0.39 | 0.49 | 99.99 | 99.99 | 0.49 | -0.08 | 0.19 | 0.52 | 1.2 | 0.09 |

## Table 8.10 (Continued)

Mean log catchability and standard e rror of ages with catchability
independent of year class strength and constant w.r.t. time

| Age | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Mean Log q | -0.551 | 0.3891 | 0.8607 | 0.9962 | 0.5897 | 0.2167 | 0.2167 | 0.2167 | 0.2167 | 0.2167 |
| S.E(Log q) | 0.4193 | 0.3497 | 0.2658 | 0.2224 | 0.3827 | 0.3007 | 0.5124 | 0.6357 | 0.7301 | 0.7258 |

Regression statistics:
Ages with q independent of year class strength and constant w.r.t. time.

| Age | Slope | t -value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 5 |  |  |  |  |  |  |  |
| 6 | 0.91 | 0.248 | 1.45 | 0.46 | 16 | 0.4 | -0.55 |
| 7 | 1.19 | -0.344 | -2.35 | 0.28 | 16 | 0.44 | 0.39 |
| 8 | 1.28 | -0.673 | -3.79 | 0.41 | 16 | 0.35 | 0.86 |
| 9 | 1.24 | -0.95 | -3.44 | 0.65 | 16 | 0.28 | 1 |
| 10 | 1.5 | -1.091 | -5.21 | 0.37 | 16 | 0.57 | 0.59 |
| 11 | 1.16 | -0.532 | -1.56 | 0.58 | 16 | 0.36 | 0.22 |
| 12 | 1.04 | -0.131 | -0.67 | 0.62 | 16 | 0.53 | 0.39 |
| 13 | 0.87 | 0.69 | 0.41 | 0.78 | 16 | 0.46 | 0.58 |
| 14 | 0.84 | 0.94 | 0.55 | 0.81 | 16 | 0.52 | 0.59 |
|  | 0.8 | 1.201 | 0.88 | 0.81 | 16 | 0.54 | 0.4 |

Fleet:FLT08: Norw.Comb.Sur

| Age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 5 | 99.99 | 99.99 | 99.99 | 99.99 | 0.26 | -0.11 | -0.31 | -0.34 |  |  |
| 6 | 99.99 | 99.99 | 99.99 | 99.99 | 0.37 | 0.2 | -0.3 | -0.03 |  |  |
| 7 | 99.99 | 99.99 | 99.99 | 99.99 | 0.41 | 0.13 | 0.22 | -0.01 |  |  |
| 8 | 99.99 | 99.99 | 99.99 | 99.99 | 0.57 | -0.28 | -0.11 | 0.32 |  |  |
| 9 | 99.99 | 99.99 | 99.99 | 99.99 | 0.06 | -0.4 | -0.63 | -0.37 |  |  |
| 10 | 99.99 | 99.99 | 99.99 | 99.99 | 0.89 | 0.45 | 0.41 | 0.47 |  |  |
| 11 | 99.99 | 99.99 | 99.99 | 99.99 | 0.19 | 0.12 | 0.15 | -0.3 |  |  |
| 12 | 99.99 | 99.99 | 99.99 | 99.99 | 0.31 | 0.48 | 0.8 | 0.84 |  |  |
| 13 | 99.99 | 99.99 | 99.99 | 99.99 | -0.36 | -1.06 | -2.9 | 0.06 |  |  |
| 14 | 99.99 | 99.99 | 99.99 | 99.99 | 0.22 | 0.08 | 0.33 | 0.23 |  |  |
| Age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 5 | 0.06 | -0.19 | 0.01 | 0.19 | -0.01 | 0.3 | 99.99 | 99.99 | 99.99 | 99.99 |
| 6 | -0.15 | 0.05 | -0.11 | 0.1 | -0.07 | 0.1 | 99.99 | 99.99 | 99.99 | 99.99 |
| 7 | -0.16 | 0.11 | 0.16 | 0.07 | -0.01 | -0.53 | 99.99 | 99.99 | 99.99 | 99.99 |
| 8 | -0.06 | -0.01 | 0.02 | -0.05 | -0.06 | -0.1 | 99.99 | 99.99 | 99.99 | 99.99 |
| 9 | 0.41 | -0.23 | 0.33 | 0.22 | -0.02 | 0.2 | 99.99 | 99.99 | 99.99 | 99.99 |
| 10 | -0.23 | 0.16 | -0.22 | -0.07 | -0.5 | -0.35 | 99.99 | 99.99 | 99.99 | 99.99 |
| 11 | -0.89 | -0.69 | -0.15 | -0.77 | -0.95 | -0.58 | 99.99 | 99.99 | 99.99 | 99.99 |
| 12 | -0.27 | -0.06 | 0.15 | -0.15 | 0.11 | -0.32 | 99.99 | 99.99 | 99.99 | 99.99 |
| 13 | -0.55 | -0.62 | -0.15 | -0.31 | -0.07 | -0.25 | 99.99 | 99.99 | 99.99 | 99.99 |
| 14 | -0.66 | -0.17 | -0.16 | -0.52 | 0.1 | -0.58 | 99.99 | 99.99 | 99.99 | 99.99 |

## Table 8.10 (Continued)

Mean $\log$ catchability and standard e rror of ages with catchability
independent of year class strength and constant w.r.t. ti me

| Age | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Mean $\log q$ | -0.2967 | 0.2253 | 0.8256 | 0.3604 | -0.1927 | 0.5791 | 0.5791 | 0.5791 | 0.5791 | 0.5791 |
| S.E(Log q) | 0.2241 | 0.1662 | 0.2595 | 0.2003 | 0.3412 | 0.4155 | 0.6655 | 0.4338 | 0.982 | 0.4103 |

Regression statistics:

Ages with q independent of year class streng th and constant w.r.t. time.

| Age | Slope | t-value | Intercept | RSquare | No Pts | Reg s.e | Mean Q |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 5 | 0.51 | 1.614 | 4.97 | 0.67 | 10 | 0.1 | -0.3 |
| 6 | 2.32 | -0.521 | -13.34 | 0.03 | 10 | 0.41 | 0.23 |
| 7 | -3.98 | -0.881 | 50.32 | 0.01 | 10 | 1.05 | 0.83 |
| 8 | 3.28 | -1.332 | -21.79 | 0.06 | 10 | 0.62 | 0.36 |
| 9 | 0.71 | 0.808 | 2.61 | 0.6 | 10 | 0.25 | -0.19 |
| 10 | 7.26 | -2.952 | -55.97 | 0.04 | 10 | 2.03 | 0.58 |
| 11 | 2.15 | -2.702 | -8.98 | 0.5 | 10 | 0.65 | 0.1 |
| 12 | 1.69 | -2.435 | -5.96 | 0.7 | 10 | 0.53 | 0.7 |
| 13 | 0.66 | 1.444 | 2.05 | 0.77 | 10 | 0.49 | 0.05 |
| 14 | 1.21 | -0.96 | -1.66 | 0.79 | 10 | 0.45 | 0.41 |
| 1 |  |  |  |  |  |  |  |

Terminal year survivor and F summaries:

Age 5 Catchability constant w.r.t. time and dependent on age

Year class $=2004$

| Fleet | Estimated | Int | Ext | Var | N | Scaled <br> Weights | Estimated |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| FLT04: Norw. Exp. CP | 1 | Survivors | s.e | s.e | Ratio |  | 0 |
| FLT07: Russ.Surv. ne | 59116 | 0 | 0 | 0 | 0 | 0 |  |
| FLT08: Norw.Comb.Sur | 1 | 0.439 | 0 | 0 | 1 | 0.56 | 0.015 |
| F shrinkage mean | 30939 | 0.5 |  |  | 0 | 0 | 0 |

Weighted prediction:

| Survivors | Int | Ext | N | Var | F |
| :--- | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 44447 | 0.33 | 0.43 | 2 | 1.302 | 0.02 |

## Table 8.10 (Continued)

Age 6 Catchability constant w.r.t. time and dependent on age

| Year class $=2003$ |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| FLT04: Norw. Exp. CP | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| FLT07: Russ.Surv. ne | 23354 | 0.281 | 0.248 | 0.88 | 2 | 0.75 | 0.044 |
| FLT08: Norw.Comb.Sur | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| F shrinkage mean | 18652 | 0.5 |  |  |  | 0.25 | 0.055 |

Weighted prediction:

| Survivors | Int | Ext | N | Var | F |
| :--- | :---: | :---: | :---: | :---: | :---: |
| at end ofyear | s.e | s.e |  | Ratio |  |
| 22076 | 0.25 | 0.17 | 3 | 0.699 | 0.047 |

Age 7 Catchability constant w.r.t. time and dependent on age
Year class $=2002$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| FLT04: Norw. Exp. CP | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| FLT07: Russ.Surv. ne | 14732 | 0.205 | 0.141 | 0.69 | 3 | 0.841 | 0.084 |
| FLT08: Norw.Comb.Sur | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| F shrinkage mean | 11828 | 0.5 |  |  |  | 0.159 | 0.103 |

Weighted prediction:

| Survivors | Int | Ext | N | Var <br> Ratio | F |
| :--- | :---: | :---: | :---: | :---: | :---: |
| at end ofyear | s.e | s.e |  |  |  |
| 14226 | 0.19 | 0.12 | 4 | 0.615 | 0.086 |

Age 8 Catchability constant w.r.t. time and dependent on age
Year class = 2001

| Fleet | Estimated Survivors | Int s.e | $\begin{aligned} & \text { Ext } \\ & \text { s.e } \end{aligned}$ | Var Ratio | N | Scaled Weights | Estimated <br> F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FLT04: Norw. Exp. CP | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| FLT07: Russ.Surv. ne | 12038 | 0.17 | 0.091 | 0.54 | 4 | 0.877 | 0.126 |
| FLT08: Norw.Comb.Sur | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| F shrinkage mean | 12779 | 0.5 |  |  |  | 0.123 | 0.119 |
| Weighted prediction: |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | Var | F |  |  |
| at end ofyear | s.e | s.e |  | Ratio |  |  |  |
| 12127 | 0.16 | 0.07 | 5 | 0.464 | 0.125 |  |  |

## Table 8.10 (Continued)

Age 9 Catchability constant w.r.t. time and dependent on age
Year class $=2000$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| FLT04: Norw. Exp. CP | 4427 | 0.616 | 0 | 0 | 1 | 0.04 | 0.141 |
| FLT07: Russ.Surv. ne | 8879 | 0.158 | 0.068 | 0.43 | 5 | 0.707 | 0.073 |
| FLT08: Norw.Comb.Sur | 11721 | 0.304 | 0 | 0 | 1 | 0.163 | 0.056 |
| F shrinkage mean | 5430 | 0.5 |  |  |  | 0.09 | 0.116 |

Weighted prediction:

| Survivors | Int | Ext | N | Var | F |
| :--- | :---: | :---: | :---: | :---: | :---: |
| at end ofyear | s.e | s.e |  | Ratio |  |
| 8649 | 0.13 | 0.1 | 8 | 0.745 | 0.075 |

Age 10 Catchability constant w.r.t. time and dependent on age
Year class $=1999$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| FLT04: Norw. Exp. CP | 4885 | 0.273 | 0.054 | 0.2 | 2 | 0.127 | 0.047 |
| FLT07: Russ.Surv. ne | 4688 | 0.143 | 0.162 | 1.13 | 6 | 0.606 | 0.049 |
| FLT08: Norw.Comb.Sur | 4786 | 0.216 | 0.055 | 0.26 | 2 | 0.201 | 0.048 |
| F shrinkage mean | 2634 | 0.5 |  |  |  | 0.066 | 0.086 |

Weighted prediction:

| Survivors | Int | Ext | N | Var | F |
| :--- | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 4555 | 0.11 | 0.1 | 11 | 0.94 | 0.051 |

Age 11 Catchability constant w.r.t. time and age (fixed at the value for age) 10 Year class $=1998$

| Fleet | Estimated | Int | Ext | Var | N | Scaled | Estimated |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| FLT04: Norw. Exp. CP | 2638 | 0.205 | 0.025 | 0.12 | 3 | 0.187 | 0.166 |
| FLT07: Russ.Surv. ne | 3393 | 0.147 | 0.089 | 0.6 | 6 | 0.504 | 0.132 |
| FLT08: Norw.Comb.Sur | 2599 | 0.178 | 0.211 | 1.19 | 3 | 0.244 | 0.168 |
| F shrinkage mean | 3442 | 0.5 |  |  |  | 0.066 | 0.13 |

Weighted prediction:

| Survivors | Int | Ext | N | Var | F |
| :--- | :---: | :---: | :---: | :---: | :---: |
| at end of year | s.e | s.e |  | Ratio |  |
| 3036 | 0.1 | 0.07 | 13 | 0.703 | 0.146 |

## Table 8.10 (Continued)

Age 12 Catchability constant w.r.t. time and age (fixed at the value for age) 10
Year class $=1997$

| Fleet | Estimated | Int | Ext | Var | N | Scaled <br> Weights | F |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | Survivors | s.e | s.e | Ratio |  | Weighted |  |
| FLT04: Norw. Exp. CP | 2266 | 0.179 | 0.157 | 0.88 | 4 | 0.219 | 0.125 |
| FLT07: Russ.Surv. ne | 2339 | 0.157 | 0.142 | 0.9 | 6 | 0.427 | 0.121 |
| FLT08: Norw.Comb.Sur | 2154 | 0.156 | 0.042 | 0.27 | 4 | 0.285 | 0.131 |
| F shrinkage mean | 1273 | 0.5 |  |  |  | 0.069 | 0.213 |

Weighted prediction:

| Survivors | Int | Ext | N | Var | F |
| :--- | :---: | :---: | :---: | :---: | :---: |
| at end of fyear | s.e | s.e |  | Ratio |  |
| 2176 | 0.1 | 0.08 | 15 | 0.81 | 0.13 |

Age 13 Catchability constant w.r.t. time and age (fixed at the value for age) 10 Year class $=1996$

| Fleet | Estimated | Int | Ext | Var | N | Scaled <br>  <br>  <br> Survivors | s.e |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Weighted prediction:

| Survivors <br> at end of year | Int | Ext | N | Var | F |
| :--- | :---: | :---: | :---: | :---: | :---: |
| 1233 | s.e | s.e |  | Ratio |  |

Age 14 Catchability constant w.r.t. time and age (fixed at the value for age) 10 Year class $=1995$

| Fleet | Estimated | Int | Ext | V ar | N | Scaled | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Survivors | s.e | s.e | Ratio |  | Weights | F |
| FLT04: Norw. Exp. CP | 827 | 0.171 | 0.144 | 0.84 | 6 | 0.233 | 0.102 |
| FLT07: Russ.Surv. ne | 1033 | 0.186 | 0.202 | 1.09 | 8 | 0.337 | 0.083 |
| FLT08: Norw.Comb.Sur | 874 | 0.144 | 0.067 | 0.47 | 6 | 0.33 | 0.097 |
| F shrinkage mean | 691 | 0.5 |  |  |  | 0.1 | 0.121 |
| Weighted prediction: |  |  |  |  |  |  |  |
| Survivors | Int | Ext | N | V ar |  |  |  |
| at end ofyear | s.e | s.e |  | Ratio |  |  |  |
| 892 | 0.1 | 0.08 | 21 | 0.836 |  |  |  |

## Table 8.11. Fishing mortality (F) at age

Run title : NEA Greenland halibut (run: 2010/1)

At 26/04/2010 18:36

Terminal Fs derived using XSA with final year \& oldest age shrinkage.

Table 8 Fishing mortality ( F ) at age

| YEAR | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| AGE |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 0.0094 | 0.0053 | 0.0032 | 0.0024 | 0.0019 | 0.0207 | 0.0139 | 0.0027 | 0.0363 | 0.0074 | 0.0378 |
| 6 | 0.0484 | 0.0255 | 0.0138 | 0.0072 | 0.0051 | 0.0484 | 0.0659 | 0.1491 | 0.151 | 0.0442 | 0.1079 |
| 7 | 0.1146 | 0.0699 | 0.0397 | 0.018 | 0.0116 | 0.0691 | 0.2864 | 0.4473 | 0.511 | 0.2369 | 0.3446 |
| 8 | 0.2531 | 0.216 | 0.1411 | 0.0891 | 0.0694 | 0.2081 | 0.6556 | 0.6021 | 0.4033 | 0.3335 | 0.3623 |
| 9 | 0.4566 | 0.2848 | 0.3476 | 0.2356 | 0.2381 | 0.2332 | 0.5603 | 0.4391 | 0.2444 | 0.2596 | 0.2744 |
| 10 | 0.7003 | 0.7254 | 0.2583 | 0.3382 | 0.3302 | 0.435 | 0.5339 | 0.4738 | 0.1999 | 0.2516 | 0.3041 |
| 11 | 0.6375 | 0.7606 | 0.5421 | 0.2684 | 0.5684 | 0.4571 | 0.4457 | 0.4037 | 0.2511 | 0.2585 | 0.3297 |
| 12 | 0.5666 | 0.8214 | 0.8585 | 0.8372 | 0.1802 | 0.3905 | 0.4362 | 0.5627 | 0.3063 | 0.3191 | 0.3545 |
| 13 | 0.4065 | 0.391 | 0.4515 | 1.0092 | 0.2945 | 0.0686 | 0.5465 | 0.7562 | 0.4414 | 0.2765 | 0.3346 |
| 14 | 0.5568 | 0.6004 | 0.4943 | 0.5409 | 0.3237 | 0.3182 | 0.5074 | 0.5302 | 0.2897 | 0.2741 | 0.3208 |
| +gp | 0.5568 | 0.6004 | 0.4943 | 0.5409 | 0.3237 | 0.3182 | 0.5074 | 0.5302 | 0.2897 | 0.2741 | 0.3208 |
| 0 FBAR 6-10 | 0.3146 | 0.2643 | 0.1601 | 0.1376 | 0.1309 | 0.1988 | 0.4204 | 0.4223 | 0.3019 | 0.2252 | 0.2787 |

Table 8 Fishing mortality ( F ) at age

| YEAR | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

AGE

| 5 | 0.041 | 0.0413 | 0.0972 | 0.1045 | 0.1292 | 0.0432 | 0.1211 | 0.077 | 0.0909 | 0.0569 | 0.0681 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 6 | 0.1211 | 0.1894 | 0.2134 | 0.2343 | 0.2394 | 0.0858 | 0.1443 | 0.1254 | 0.1426 | 0.3077 | 0.2402 |
| 7 | 0.4196 | 0.4664 | 0.4174 | 0.4302 | 0.2654 | 0.1813 | 0.1929 | 0.1279 | 0.2136 | 0.3858 | 0.3424 |
| 8 | 0.3817 | 0.625 | 0.3556 | 0.4139 | 0.2072 | 0.1908 | 0.1386 | 0.1692 | 0.3342 | 0.3421 | 0.2913 |
| 9 | 0.3557 | 0.4999 | 0.3926 | 0.3518 | 0.1331 | 0.229 | 0.0923 | 0.3235 | 0.307 | 0.2414 | 0.2714 |
| 10 | 0.4017 | 0.3508 | 0.3248 | 0.3979 | 0.1093 | 0.1721 | 0.153 | 0.3451 | 0.4542 | 0.4057 | 0.3698 |
| 11 | 0.5023 | 0.3823 | 0.4845 | 0.4735 | 0.1956 | 0.2421 | 0.2515 | 0.4452 | 0.3166 | 0.3966 | 0.3562 |
| 12 | 0.5617 | 0.6828 | 0.708 | 0.3548 | 0.2022 | 0.2655 | 0.27 | 0.4246 | 0.4772 | 0.2311 | 0.4171 |
| 13 | 0.5354 | 0.5073 | 0.8178 | 0.6669 | 0.1237 | 0.3001 | 0.6798 | 0.3669 | 0.3601 | 0.2863 | 0.1544 |
| 14 | 0.4739 | 0.4873 | 0.5488 | 0.4513 | 0.1532 | 0.2426 | 0.2905 | 0.3829 | 0.3848 | 0.3135 | 0.3151 |
|  | +gp | 0.4739 | 0.4873 | 0.5488 | 0.4513 | 0.1532 | 0.2426 | 0.2905 | 0.3829 | 0.3848 | 0.3135 |
| 0 | FBAR $6-10$ | 0.336 | 0.4263 | 0.3408 | 0.3656 | 0.1909 | 0.1718 | 0.1442 | 0.2182 | 0.2903 | 0.3365 |
| 0.303 |  |  |  |  |  |  |  |  |  |  |  |

## Table 8.11 (Continued)

Terminal Fs derived using XSA with final year \& oldest age shrinkage.

Table 8 Fishing mortality ( F ) at age

| YEAR | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

AGE

| 5 | 0.095 | 0.0695 | 0.0433 | 0.114 | 0.1721 | 0.3283 | 0.1181 | 0.0983 | 0.0373 | 0.051 | 0.0601 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 6 | 0.2536 | 0.2303 | 0.1926 | 0.2912 | 0.4281 | 0.5053 | 0.1778 | 0.1558 | 0.0775 | 0.0719 | 0.1623 |
| 7 | 0.3534 | 0.445 | 0.3827 | 0.4387 | 0.5267 | 0.8372 | 0.2357 | 0.3628 | 0.2555 | 0.2577 | 0.4027 |
| 8 | 0.3335 | 0.3813 | 0.4815 | 0.3366 | 0.4127 | 0.5254 | 0.2891 | 0.39 | 0.2986 | 0.3112 | 0.3331 |
| 9 | 0.3374 | 0.256 | 0.4543 | 0.321 | 0.4194 | 0.3809 | 0.131 | 0.0727 | 0.1668 | 0.2237 | 0.119 |
| 10 | 0.4645 | 0.4189 | 0.4905 | 0.1982 | 0.3202 | 1.0123 | 0.3761 | 0.5811 | 0.5134 | 0.6811 | 0.6367 |
| 11 | 0.3089 | 0.2869 | 0.4325 | 0.2169 | 0.2375 | 1.1399 | 0.3538 | 0.4986 | 0.4946 | 0.8203 | 0.5478 |
| 12 | 0.4329 | 0.1747 | 0.4062 | 0.1799 | 0.4904 | 1.5795 | 0.6517 | 0.4745 | 0.8168 | 1.0875 | 0.5497 |
| 13 | 0.7343 | 0.3144 | 0.1578 | 0.2525 | 0.0794 | 0.5353 | 0.7454 | 0.296 | 0.564 | 1.155 | 0.1968 |
| 14 | 0.458 | 0.2913 | 0.3901 | 0.2345 | 0.3107 | 0.9371 | 0.5338 | 0.394 | 0.568 | 0.8213 | 0.4476 |
|  | +gp | 0.458 | 0.2913 | 0.3901 | 0.2345 | 0.3107 | 0.9371 | 0.5338 | 0.394 | 0.568 | 0.8213 |
| 0.4476 |  |  |  |  |  |  |  |  |  |  |  |
| FBAR 6-10 | 0.3485 | 0.3463 | 0.4003 | 0.3171 | 0.4214 | 0.6522 | 0.2419 | 0.3125 | 0.2623 | 0.3091 | 0.3308 |

Table 8 Fishing mortality (F) at age

| YEAR | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


|  | AGE |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 5 |  | 0.0167 | 0.0192 | 0.0265 | 0.0223 | 0.0263 | 0.0149 | 0.0225 | 0.0158 | 0.0171 | 0.0203 | 0.0412 |
| 6 |  | 0.0662 | 0.0685 | 0.1273 | 0.0544 | 0.0977 | 0.067 | 0.0668 | 0.0701 | 0.0709 | 0.0651 | 0.0719 |
| 7 |  | 0.1973 | 0.2395 | 0.3451 | 0.1738 | 0.2315 | 0.1399 | 0.1739 | 0.2002 | 0.1778 | 0.18 | 0.1366 |
| 8 |  | 0.1563 | 0.2277 | 0.2623 | 0.1951 | 0.1645 | 0.1361 | 0.1162 | 0.1841 | 0.2186 | 0.2016 | 0.1343 |
| 9 |  | 0.1167 | 0.115 | 0.2051 | 0.1544 | 0.1447 | 0.1042 | 0.1153 | 0.1836 | 0.1718 | 0.1818 | 0.1087 |
| 10 |  | 0.6177 | 0.492 | 0.7176 | 0.4804 | 0.3532 | 0.2901 | 0.2374 | 0.1922 | 0.2402 | 0.1753 | 0.1028 |
| 11 |  | 0.4538 | 0.3373 | 0.3619 | 0.346 | 0.3315 | 0.2785 | 0.205 | 0.2245 | 0.2168 | 0.1677 | 0.1476 |
| 12 |  | 0.6364 | 0.4345 | 0.6274 | 0.5798 | 0.4959 | 0.3077 | 0.2308 | 0.3691 | 0.3139 | 0.288 | 0.2268 |
| 13 |  | 0.1234 | 0.1091 | 0.4265 | 0.2353 | 0.3828 | 0.3074 | 0.1133 | 0.2144 | 0.3876 | 0.3871 | 0.1912 |
| 14 |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | 0.4419 | 0.2822 | 0.47 | 0.4174 | 0.415 | 0.3258 | 0.414 | 0.3292 | 0.4166 | 0.3649 | 0.2221 |
|  | +gp | 0.4419 | 0.2822 | 0.47 | 0.4174 | 0.415 | 0.3258 | 0.414 | 0.3292 | 0.4166 | 0.3649 | 0.2221 |
| FBAR 6-10 | 0.2308 | 0.2285 | 0.3315 | 0.2116 | 0.1983 | 0.1475 | 0.1419 | 0.166 | 0.1759 | 0.1607 | 0.1109 |  |


| AGE |  |  |  |
| :---: | :--- | :--- | :--- |
|  | 0.0154 | 0.0198 | 0.0255 |
|  | 0.0379 | 0.0467 | 0.0522 |
|  | 0.0691 | 0.0864 | 0.0974 |
|  | 0.1038 | 0.1254 | 0.1212 |
|  | 0.1237 | 0.0747 | 0.1024 |
|  | 0.0689 | 0.0506 | 0.0741 |
|  | 0.1112 | 0.1459 | 0.1349 |
|  | 0.1969 | 0.1298 | 0.1845 |
|  | 0.245 | 0.2035 | 0.2133 |
|  | 0.2307 | 0.0952 | 0.1827 |
| +gp | 0.2307 | 0.0952 |  |
| FBAR 6-10 | 0.0807 | 0.0768 |  |

Table 8.12. Stock number at age (start of year) Numbers* ${ }^{*} 0^{* *}-3$
Run title : NEA Greenland halibut (run: 2010/1)

| At 26/04/2010 18:36 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Table 10 Stock number at age (start of year) |  |  |  |  |  | Numbers*10**-3 |  |  |  |  |  |  |
|  | YEAR | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |
| 5 |  | 42840 | 51686 | 57829 | 70444 | 64281 | 55933 | 41113 | 31553 | 33558 | 31064 | 26648 |
| 6 |  | 33792 | 36528 | 44252 | 49616 | 60487 | 55222 | 47154 | 34898 | 27083 | 27855 | 26540 |
| 7 |  | 27961 | 27712 | 30649 | 37566 | 42397 | 51799 | 45285 | 37996 | 25876 | 20044 | 22939 |
| 8 |  | 27353 | 21461 | 22243 | 25353 | 31755 | 36072 | 41608 | 29269 | 20910 | 13360 | 13613 |
| 9 |  | 14559 | 18279 | 14883 | 16626 | 19961 | 25499 | 25214 | 18591 | 13796 | 12025 | 8238 |
| 10 |  | 8521 | 7938 | 11834 | 9049 | 11307 | 13541 | 17381 | 12393 | 10314 | 9300 | 7983 |
| 11 |  | 4237 | 3641 | 3307 | 7867 | 5554 | 6995 | 7544 | 8771 | 6641 | 7269 | 6224 |
| 12 |  | 2537 | 1928 | 1465 | 1656 | 5177 | 2707 | 3812 | 4158 | 5042 | 4447 | 4832 |
| 13 |  | 1175 | 1239 | 730 | 534 | 617 | 3721 | 1577 | 2121 | 2039 | 3195 | 2782 |
| 14 |  | 634 | 673 | 721 | 400 | 168 | 395 | 2990 | 786 | 857 | 1129 | 2085 |
|  | +gp | 190 | 118 | 77 | 49 | 27 | 118 | 756 | 372 | 341 | 564 | 844 |
| 0 | TOTAL | 163799 | 171203 | 187989 | 219159 | 241730 | 252002 | 234434 | 180908 | 146458 | 130252 | 122729 |
|  | Table 10 Stock number at age (start of year) |  |  |  |  | Numbers*10**3 |  |  |  |  |  |  |
|  | YEAR | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |
| 5 |  | 22548 | 22108 | 23711 | 20604 | 19729 | 18655 | 17916 | 18968 | 19148 | 17841 | 19960 |
| 6 |  | 22086 | 18628 | 18258 | 18518 | 15974 | 14923 | 15378 | 13661 | 15116 | 15049 | 14507 |
| 7 |  | 20506 | 16841 | 13266 | 12695 | 12609 | 10822 | 11788 | 11458 | 10372 | 11281 | 9522 |
| 8 |  | 13988 | 11601 | 9092 | 7522 | 7106 | 8323 | 7770 | 8366 | 8678 | 7210 | 6602 |
| 9 |  | 8156 | 8219 | 5345 | 5484 | 4280 | 4972 | 5919 | 5822 | 6080 | 5347 | 4408 |
| 10 |  | 5389 | 4919 | 4291 | 3106 | 3320 | 3225 | 3403 | 4646 | 3626 | 3850 | 3616 |
| 11 |  | 5069 | 3104 | 2981 | 2669 | 1796 | 2562 | 2337 | 2514 | 2832 | 1982 | 2209 |
| 12 |  | 3853 | 2640 | 1823 | 1580 | 1431 | 1271 | 1731 | 1564 | 1386 | 1776 | 1147 |
| 13 |  | 2917 | 1891 | 1148 | 773 | 954 | 1006 | 839 | 1137 | 880 | 740 | 1213 |
| 14 |  | 1713 | 1470 | 980 | 436 | 341 | 726 | 641 | 366 | 678 | 529 | 479 |
|  | +gp | 1044 | 993 | 456 | 330 | 386 | 389 | 264 | 155 | 215 | 283 | 250 |
| 0 | TOTAL | 107269 | 92414 | 81351 | 73719 | 67927 | 66873 | 67988 | 68657 | 69011 | 65888 | 63912 |
|  | Table 10 | Stock number at age (start of year) |  |  |  | Numbers*10**-3 |  |  |  |  |  |  |
|  | YEAR | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |
| 5 |  | 19896 | 19464 | 23050 | 20808 | 14584 | 12756 | 10626 | 13125 | 18579 | 18325 | 19117 |
| 6 |  | 16049 | 15573 | 15628 | 18998 | 15980 | 10568 | 7907 | 8127 | 10239 | 15406 | 14988 |
| 7 |  | 9820 | 10719 | 10647 | 11095 | 12220 | 8964 | 5488 | 5697 | 5986 | 8156 | 12340 |
| 8 |  | 5820 | 5936 | 5912 | 6250 | 6158 | 6212 | 3340 | 3732 | 3411 | 3991 | 5425 |
| 9 |  | 4246 | 3589 | 3489 | 3144 | 3842 | 3508 | 3162 | 2153 | 2175 | 2178 | 2516 |
| 10 |  | 2892 | 2608 | 2391 | 1907 | 1963 | 2174 | 2063 | 2387 | 1723 | 1584 | 1499 |
| 11 |  | 2150 | 1565 | 1477 | 1260 | 1346 | 1227 | 680 | 1219 | 1149 | 888 | 690 |
| 12 |  | 1331 | 1359 | 1011 | 825 | 873 | 914 | 338 | 411 | 637 | 603 | 336 |
| 13 |  | 651 | 743 | 982 | 580 | 593 | 460 | 162 | 152 | 220 | 242 | 175 |
| 14 |  | 895 | 269 | 467 | 722 | 388 | 471 | 232 | 66 | 97 | 108 | 66 |
|  | +gp | 696 | 30 | 156 | 144 | 177 | 905 | 145 | 13 | 7 | 15 | 3 |
| 0 | TOT AL | 64446 | 61854 | 65210 | 65732 | 58124 | 48159 | 34142 | 37082 | 44225 | 51497 | 57155 |

Table 8.12 (Continued)

|  | Table 10 <br> YEAR | Stock numberatage (start of year) |  |  |  | Numbers* $10^{* *}-3$ |  |  | 2004 | 2005 | 2006 | 2007 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1997 | 1998 | 1999 | 2000 | 02001 | 2002 | 2003 |  |  |  |  |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |
| 5 |  | 21470 | 20340 | 17846 | 18538 | 818284 | 20155 | 19224 | 19915 | 27240 | 29430 | 26328 |
| 6 |  | 15495 | 18173 | 17173 | 14958 | 815603 | 15328 | 17091 | 16178 | 16872 | 23048 | 24823 |
| 7 |  | 10968 | 12482 | 14606 | 13014 | 412193 | 12180 | 12338 | 13759 | 12982 | 13528 | 18588 |
| 8 |  | 7100 | 7750 | 8455 | 8903 | 39414 | 8326 | 9115 | 8925 | 9694 | 9353 | 9726 |
| 9 |  | 3347 | 5227 | 5312 | 5598 | $8 \quad 6304$ | 6874 | 6254 | 6984 | 6390 | 6705 | 6581 |
| 10 |  | 1923 | 2563 | 4010 | 3724 | $4 \quad 4129$ | 4695 | 5331 | 4796 | 5003 | 4632 | 4812 |
| 11 |  | 683 | 892 | 1349 | 1684 | 41983 | 2496 | 3023 | 3619 | 3407 | 3387 | 3345 |
| 12 |  | 343 | 373 | 548 | 808 | 81026 | 1225 | 1626 | 2120 | 2489 | 2361 | 2465 |
| 13 |  | 167 | 156 | 208 | 252 | 2390 | 538 | 775 | 1111 | 1261 | 1565 | 1523 |
| 14 |  | 124 | 127 | 121 | 117 | 7171 | 229 | 340 | 596 | 772 | 737 | 915 |
|  | +gp | 3 | 65 | 20 | 44 | 438 | 162 | 82 | 401 | 1031 | 806 | 1208 |
| 0 | TOT AL | 61621 | 68149 | 69649 | 67640 | 069535 | 72207 | 75200 | 78405 | 87141 | 95551 | 100313 |
|  | Table 10 | Stock numberatage (start of year) |  |  |  | Numbers* ${ }^{*} 0^{* *}-3$ |  |  |  |  |  |  |
|  | YEAR | 2008 | 2009 | 2010 |  | GMST 64-** | AMS | 64-** |  |  |  |  |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |
| 5 |  | 31706 | 52671 | 0 |  | 23733 |  | 26301 |  |  |  |  |
| 6 |  | 21745 | 26873 | 44447 |  | 19519 |  | 22039 |  |  |  |  |
| 7 |  | 19882 | 18019 | 22076 |  | 14745 |  | 17254 |  |  |  |  |
| 8 |  | 13956 | 15971 | 14226 |  | 9747 |  | 12096 |  |  |  |  |
| 9 |  | 7319 | 10827 | 12127 |  | 6360 |  | 7984 |  |  |  |  |
| 10 |  | 5081 | 5566 | 8649 |  | 4289 |  | 5267 |  |  |  |  |
| 11 |  | 3737 | 4082 | 4555 |  | 2480 |  | 3128 |  |  |  |  |
| 12 |  | 2484 | 2878 | 3036 |  | 1439 |  | 1863 |  |  |  |  |
| 13 |  | 1691 | 1756 | 2176 |  | 752 |  | 1049 |  |  |  |  |
| 14 |  | 1083 | 1139 | 1233 |  | 428 |  | 619 |  |  |  |  |
|  | +gp | 1287 | 1362 | 1957 |  |  |  |  |  |  |  |  |
| 0 | TOT AL | 109972 | 141146 | 114483 |  |  |  |  |  |  |  |  |

Table 8.13. Stock biomass at age (start of year) Tonnes


## Table 8.13 (Continued)

| Table 12 Stock biomass atage (start of year) |  |  |  |  | Tonnes |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 5 | 16532 | 14848 | 12492 | 14089 | 13530 | 13907 | 13745 | 15335 | 18224 | 18747 | 16481 | 22036 | 29865 |
| 6 | 14565 | 16901 | 16315 | 14510 | 16071 | 14408 | 17945 | 17715 | 16062 | 19821 | 22415 | 19984 | 21552 |
| 7 | 14039 | 16226 | 18550 | 17308 | 16948 | 16565 | 17619 | 20611 | 16954 | 15543 | 24406 | 27020 | 19299 |
| 8 | 11644 | 12477 | 13106 | 14511 | 16475 | 13987 | 15933 | 16984 | 16024 | 14310 | 16397 | 24506 | 23493 |
| 9 | 6927 | 11081 | 10624 | 11813 | 14437 | 14985 | 14496 | 17203 | 13617 | 14229 | 15274 | 16328 | 20875 |
| 10 | 4980 | 6587 | 9865 | 9720 | 11066 | 12583 | 13940 | 13310 | 12728 | 12144 | 12285 | 12082 | 12335 |
| 11 | 2252 | 2900 | 4343 | 5642 | 6602 | 7964 | 9200 | 11320 | 9702 | 9141 | 9785 | 10670 | 10735 |
| 12 | 1377 | 1459 | 2110 | 3210 | 4020 | 4765 | 6008 | 8074 | 8297 | 7825 | 7861 | 8024 | 8870 |
| 13 | 807 | 767 | 959 | 1252 | 1874 | 2398 | 3539 | 4769 | 4710 | 6256 | 5708 | 5997 | 6657 |
| 14 | 736 | 720 | 705 | 680 | 996 | 1201 | 1894 | 3248 | 3384 | 3420 | 4151 | 4240 | 5159 |
| +gp | 19 | 321 | 120 | 317 | 280 | 1024 | 523 | 2551 | 5972 | 5438 | 10963 | 9594 | 9627 |
| 0 TOT ALBIO | 73879 | 84287 | 89189 | 93051 | 102300 | 103785 | 114843 | 131119 | 125675 | 126875 | 145727 | 160481 | 168467 |

Table 8.14. Spawning stock biomass at age (spawning time)
Tonnes
Run title : NEA Greenland halibut (run: 2010/1) At 26/04/2010 18: 36

|  | Table 13 Sp | Spawning stock biomass a tage (spawning time) |  |  |  |  | Tonnes |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |
| 5 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 |  | 649 | 701 | 850 | 968 | 1198 | 1060 | 1043 | 772 | 599 | 616 | 587 |
| 7 |  | 755 | 748 | 837 | 1048 | 1221 | 1414 | 1466 | 1230 | 838 | 649 | 743 |
| 8 |  | 6893 | 5498 | 5792 | 6762 | 8736 | 9469 | 12416 | 8734 | 6240 | 3987 | 4062 |
| 9 |  | 15900 | 20330 | 16952 | 19048 | 23270 | 28018 | 31219 | 23019 | 17082 | 14888 | 10200 |
| 10 |  | 16562 | 15223 | 22593 | 17121 | 21295 | 26202 | 34096 | 24311 | 20233 | 18244 | 15660 |
| 11 |  | 12914 | 10704 | 9529 | 21895 | 15185 | 20496 | 21344 | 24816 | 18789 | 20566 | 17611 |
| 12 |  | 9298 | 6594 | 4866 | 5354 | 16185 | 9631 | 12129 | 13231 | 16044 | 14150 | 15374 |
| 13 |  | 5368 | 5452 | 3196 | 2281 | 2634 | 17415 | 6786 | 9127 | 8773 | 13747 | 11971 |
| 14 |  | 3175 | 3306 | 3491 | 1952 | 838 | 2128 | 14746 | 3875 | 4226 | 5565 | 10284 |
|  | +gp | 1131 | 697 | 452 | 282 | 163 | 707 | 4378 | 2171 | 2060 | 3388 | 5035 |
| 0 | TOTSPBIO | 72644 | 69254 | 68558 | 76710 | 90724 | 116541 | 139622 | 111285 | 94884 | 95800 | 91525 |
|  | Table 13 Sp | Spawning | ck biom | ss atag | pawni | time) | Tonnes |  |  |  |  |  |
|  | YEAR | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |
| 5 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 |  | 488 | 412 | 404 | 409 | 575 | 390 | 388 | 344 | 472 | 578 | 516 |
| 7 |  | 664 | 545 | 429 | 411 | 567 | 370 | 407 | 354 | 417 | 399 | 457 |
| 8 |  | 4174 | 3462 | 2713 | 2245 | 2686 | 2566 | 2546 | 2301 | 2452 | 1986 | 2321 |
| 9 |  | 10098 | 10177 | 6617 | 6790 | 6309 | 5923 | 8091 | 6788 | 7187 | 7535 | 7421 |
| 10 |  | 10572 | 9648 | 8418 | 6094 | 7424 | 6384 | 7522 | 8950 | 8117 | 9171 | 9773 |
| 11 |  | 14342 | 8782 | 8434 | 7552 | 5281 | 6688 | 6824 | 6824 | 8944 | 6651 | 7755 |
| 12 |  | 12259 | 8402 | 5800 | 5029 | 4908 | 3795 | 5818 | 5165 | 5733 | 6561 | 4486 |
| 13 |  | 12553 | 8137 | 4941 | 3326 | 3911 | 3388 | 3465 | 4913 | 4147 | 2954 | 5434 |
| 14 |  | 8449 | 7248 | 4833 | 2151 | 1639 | 3109 | 3002 | 1958 | 4124 | 2300 | 2034 |
|  | +gp | 6169 | 5884 | 2747 | 1950 | 2383 | 2078 | 1583 | 904 | 1315 | 1280 | 1206 |
| 0 | TOTSPBIO | - 79769 | 62698 | 45336 | 35957 | 35684 | 34692 | 39645 | 38501 | 42906 | 39415 | 41404 |
|  | Table 13 Spar | Spawning | ock biom | ass atag | (spawni | g time) | Tonnes |  |  |  |  |  |
|  | YEAR | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |
| 5 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 104 | 134 | 134 | 0 |
| 6 |  | 443 | 156 | 150 | 196 | 169 | 111 | 77 | 83 | 96 | 145 | 0 |
| 7 |  | 377 | 271 | 133 | 293 | 315 | 495 | 418 | 615 | 532 | 816 | 1132 |
| 8 |  | 2654 | 2198 | 2019 | 2025 | 1780 | 1631 | 1646 | 2245 | 1995 | 2014 | 2360 |
| 9 |  | 7792 | 5879 | 4002 | 3728 | 4115 | 4168 | 4612 | 3602 | 3286 | 2640 | 3269 |
| 10 |  | 8186 | 7000 | 6027 | 4775 | 3945 | 4352 | 4542 | 5290 | 3518 | 3141 | 3417 |
| 11 |  | 7130 | 5272 | 4476 | 3780 | 3516 | 3046 | 1840 | 3679 | 3243 | 2514 | 2202 |
| 12 |  | 4854 | 5060 | 3626 | 3340 | 3012 | 2997 | 1212 | 1657 | 1971 | 1793 | 1271 |
| 13 |  | 2603 | 3389 | 4171 | 2486 | 2206 | 1790 | 621 | 770 | 845 | 928 | 831 |
| 14 |  | 3740 | 1344 | 1956 | 3248 | 1585 | 2065 | 986 | 418 | 480 | 537 | 410 |
|  | +gp | 3152 | 177 | 696 | 679 | 799 | 4787 | 695 | 117 | 49 | 124 | 18 |
| 0 | TOTSPBIO | - 40930 | 30746 | 27256 | 24550 | 21442 | 25442 | 16648 | 18581 | 16149 | 14786 | 14910 |

Table 8.14 (Continued)

|  | Table 13 Sp | Spawning st | biomas | age (sp | ning tim | Ton |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | YEAR | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |
| 5 |  | 0 | 0 | 0 | 0 | 135 | 139 | 137 | 0 | 0 | 0 | 0 |
| 6 |  | 0 | 0 | 0 | 145 | 482 | 432 | 359 | 177 | 161 | 198 | 224 |
| 7 |  | 983 | 649 | 371 | 519 | 1017 | 1656 | 1938 | 1649 | 848 | 777 | 976 |
| 8 |  | 2445 | 1248 | 917 | 1451 | 3130 | 4336 | 5417 | 4755 | 3525 | 2576 | 2132 |
| 9 |  | 3672 | 4987 | 3506 | 4371 | 7074 | 9890 | 10437 | 11354 | 7762 | 7114 | 5193 |
| 10 |  | 4233 | 5402 | 6511 | 6124 | 7193 | 9940 | 12267 | 12112 | 11200 | 8987 | 6511 |
| 11 |  | 2117 | 2668 | 3735 | 4908 | 5546 | 7247 | 8464 | 10641 | 8829 | 7770 | 6458 |
| 12 |  | 1295 | 1459 | 2089 | 3081 | 3859 | 4574 | 5828 | 7751 | 7882 | 7277 | 6289 |
| 13 |  | 807 | 767 | 959 | 1252 | 1874 | 2374 | 3468 | 4673 | 4663 | 6131 | 4909 |
| 14 |  | 736 | 720 | 705 | 680 | 996 | 1201 | 1856 | 3183 | 3316 | 3386 | 3985 |
|  | +gp | 19 | 321 | 120 | 317 | 280 | 1024 | 523 | 2551 | 5972 | 5438 | 10854 |
| 0 | TOTSPBIO | 16306 | 18220 | 18913 | 22849 | 31587 | 42814 | 50696 | 58847 | 54158 | 49654 | 47531 |
|  | Table 13 | Spawning st | biomas | tage (sp | ning tim | Tonn |  |  |  |  |  |  |
|  | YEAR | 2008 | 2009 |  |  |  |  |  |  |  |  |  |
|  | AGE |  |  |  |  |  |  |  |  |  |  |  |
| 5 |  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 6 |  | 200 | 431 |  |  |  |  |  |  |  |  |  |
| 7 |  | 811 | 579 |  |  |  |  |  |  |  |  |  |
| 8 |  | 1715 | 2819 |  |  |  |  |  |  |  |  |  |
| 9 |  | 3919 | 8767 |  |  |  |  |  |  |  |  |  |
| 10 |  | 4349 | 7278 |  |  |  |  |  |  |  |  |  |
| 11 |  | 6188 | 8480 |  |  |  |  |  |  |  |  |  |
| 12 |  | 5858 | 7628 |  |  |  |  |  |  |  |  |  |
| 13 |  | 4918 | 6324 |  |  |  |  |  |  |  |  |  |
| 14 |  | 4070 | 5159 |  |  |  |  |  |  |  |  |  |
|  | +gp | 9498 | 9627 |  |  |  |  |  |  |  |  |  |
| 0 | TOTSPBIO | 41526 | 57093 |  |  |  |  |  |  |  |  |  |

Table 8.15. Summary (without SOP correction)

| Run title : NEA Greenland halibut (run: 2010/1) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| At 26/04/2010 18:36 |  |  |  |  |  |  |
|  | RECRUITS | TOTALBIO | TOTSPBIO | LANDINGS | YIELD/SSB | FBAR 6-10 |
| Age 5 |  |  |  |  |  |  |
| 1964 | 42840 | 172936 | 72644 | 40391 | 0.556 | 0.3146 |
| 1965 | 51686 | 171360 | 69254 | 34751 | 0.5018 | 0.2643 |
| 1966 | 57829 | 181481 | 68558 | 26321 | 0.3839 | 0.1601 |
| 1967 | 70444 | 209629 | 76710 | 24267 | 0.3163 | 0.1376 |
| 1968 | 64281 | 244357 | 90724 | 26168 | 0.2884 | 0.1309 |
| 1969 | 55933 | 274338 | 116541 | 43789 | 0.3757 | 0.1988 |
| 1970 | 41113 | 312358 | 139622 | 89484 | 0.6409 | 0.4204 |
| 1971 | 31553 | 242821 | 111285 | 79034 | 0.7102 | 0.4223 |
| 1972 | 33558 | 196247 | 94884 | 43055 | 0.4538 | 0.3019 |
| 1973 | 31064 | 180315 | 95800 | 29938 | 0.3125 | 0.2252 |
| 1974 | 26648 | 173145 | 91525 | 37763 | 0.4126 | 0.2787 |
| 1975 | 22548 | 152745 | 79769 | 38172 | 0.4785 | 0.336 |
| 1976 | 22108 | 126133 | 62698 | 36074 | 0.5754 | 0.4263 |
| 1977 | 23711 | 100845 | 45336 | 28827 | 0.6359 | 0.3408 |
| 1978 | 20604 | 87203 | 35957 | 24617 | 0.6846 | 0.3656 |
| 1979 | 19729 | 105009 | 35684 | 17312 | 0.4852 | 0.1909 |
| 1980 | 18655 | 86210 | 34692 | 13284 | 0.3829 | 0.1718 |
| 1981 | 17916 | 92194 | 39645 | 15018 | 0.3788 | 0.1442 |
| 1982 | 18968 | 87871 | 38501 | 16789 | 0.4361 | 0.2182 |
| 1983 | 19148 | 104233 | 42906 | 22147 | 0.5162 | 0.2903 |
| 1984 | 17841 | 93518 | 39415 | 21883 | 0.5552 | 0.3365 |
| 1985 | 19960 | 92643 | 41404 | 19945 | 0.4817 | 0.303 |
| 1986 | 19896 | 91902 | 40930 | 22875 | 0.5589 | 0.3485 |
| 1987 | 19464 | 85288 | 30746 | 19112 | 0.6216 | 0.3463 |
| 1988 | 23050 | 85033 | 27256 | 19587 | 0.7186 | 0.4003 |
| 1989 | 20808 | 88855 | 24550 | 20138 | 0.8203 | 0.3171 |
| 1990 | 14584 | 78185 | 21442 | 23183 | 1.0812 | 0.4214 |
| 1991 | 12756 | 72592 | 25442 | 33320 | 1.3097 | 0.6522 |
| 1992 | 10626 | 45641 | 16648 | 8602 | 0.5167 | 0.2419 |
| 1993 | 13125 | 52285 | 18581 | 11933 | 0.6422 | 0.3125 |
| 1994 | 18579 | 52462 | 16149 | 9226 | 0.5713 | 0.2623 |
| 1995 | 18325 | 59750 | 14786 | 11734 | 0.7936 | 0.3091 |
| 1996 | 19117 | 69264 | 14910 | 14347 | 0.9622 | 0.3308 |
| 1997 | 21470 | 73879 | 16306 | 9410 | 0.5771 | 0.2308 |
| 1998 | 20340 | 84287 | 18220 | 11893 | 0.6528 | 0.2285 |
| 1999 | 17846 | 89189 | 18913 | 19517 | 1.0319 | 0.3315 |
| 2000 | 18538 | 93051 | 22849 | 14437 | 0.6319 | 0.2116 |
| 2001 | 18284 | 102300 | 31587 | 16307 | 0.5163 | 0.1983 |
| 2002 | 20155 | 103785 | 42814 | 13161 | 0.3074 | 0.1475 |
| 2003 | 19224 | 114843 | 50696 | 13578 | 0.2678 | 0.1419 |
| 2004 | 19915 | 131119 | 58847 | 18800 | 0.3195 | 0.166 |
| 2005 | 27240 | 125675 | 54158 | 18834 | 0.3478 | 0.1759 |
| 2006 | 29430 | 126875 | 49654 | 17897 | 0.3604 | 0.1607 |
| 2007 | 26328 | 145727 | 47531 | 15237 | 0.3206 | 0.1109 |
| 2008 | 31706 | 160481 | 41526 | 13778 | 0.3318 | 0.0807 |
| 2009 | 52671 | 168467 | 57093 | 12407 | 0.2173 | 0.0768 |
| Arith. |  |  |  |  |  |  |
| Mean | 26992 | 125838 | 49678 | 24312 | 0.5444 | 0.2648 |
| 0 Units | (Thousands) | (Tonnes) | (Tonnes) | (Tonnes) |  |  |

Table E1. GREENLAND HALIBUT in Sub-area I and II. Norwegian bottom trawl survey indices (numbers in thousands) in the Svalbard area (Division IIb).

| Year |  | $\begin{gathered} \text { Fish }<20 \\ \mathrm{~cm}^{2} \end{gathered}$ | Age |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | $9+$ |  |
| 1981 |  |  | 2.1 |  |  |  |  |  |  |  |  |  | 20100 |
| 1982 |  | 0.7 |  |  |  | No ag | data |  |  |  |  | 2600 |
| 1983 |  | 5.9 |  |  |  |  |  |  |  |  |  | 26690 |
| 1984 |  | 3.2 | 550 | 3042 | 2924 | 8573 | 6847 | 5657 | 4345 | 2796 | 1896 | 36630 |
| 1985 |  | 1.6 | 884 | 3921 | 4294 | 6674 | 8793 | 8622 | 3920 | 1817 | 525 | 39450 |
| 1986 |  | 0.1 | 49 | 1005 | 1967 | 7314 | 4671 | 1754 | 2301 | 372 | 37 | 19470 |
| 1987 |  | 1 | 630 | 1014 | 3076 | 4409 | 4786 | 3141 | 964 | 364 | 116 | 18500 |
| 1988 |  | 2.5 | 818 | 4298 | 6191 | 6696 | 12289 | 2396 | 6015 | 338 | 1277 | 40318 |
| 1989 | 1 | 1.4 | 712 | 3232 | 8158 | 7493 | 7069 | 2374 | 1753 | 353 | 744 | 31888 |
| 1990 | 1 | 0.4 | 115 | 336 | 5050 | 7130 | 7730 | 4490 | 2330 | 918 | 544 | 28643 |
| 1991 | 1 | 0.1 | 71 | 877 | 3080 | 6720 | 9270 | 5450 | 2800 | 1660 | 524 | 30452 |
| 1992 | 1 | + | 33 | 30 | 338 | 1190 | 3520 | 4420 | 2280 | 1280 | 474 | 13565 |
| 1993 | 1 | + | 25 | 60 | 51 | 1049 | 2369 | 2056 | 2772 | 1114 | 665 | 10161 |
| 1994 | 1 | + | 4 | 238 | 296 | 652 | 2775 | 2371 | 2593 | 531 | 844 | 10304 |
| 1995 | 1 | 0.1 | 76 | + | + | 322 | 886 | 1200 | 1950 | 487 | 497 | 5418 |
| 1996 | 1 | 0.4 | 410 | 61 | 104 | 171 | 881 | 2052 | 2587 | 862 | 976 | 8104 |
| 1997 | 1 | 0.4 | 268 | 484 | 21 | 65 | 284 | 2089 | 2143 | 379 | 295 | 6028 |
| 1998 | 1 | 2.5 | 1999 | 2351 | 2715 | 493 | 609 | 2192 | 2814 | 1252 | 822 | 15247 |
| 1999 | 1 | 1.3 | 126 | + | 995 | 1789 | 415 | 709 | 2501 | 507 | 674 | 7716 |
| 2000 | 1 | 2 | 2009 | 540 | 323 | 1347 | 2135 | 2634 | 1784 | 1197 | 530 | 12499 |
| 2001 | 1 | 4.3 | 4258 | 1235 | 873 | 1506 | 2456 | 1718 | 1504 | 558 | 1079 | 15187 |
| 2002 | 1 | 2.3 | 1435 | 2019 | 1176 | 2437 | 3413 | 2685 | 3304 | 847 | 2229 | 19545 |
| 2003 | 1 | 0.8 | 410 | 638 | 901 | 2937 | 2630 | 3146 | 2602 | 452 | 684 | 14400 |

${ }^{1}$ New standard trawl equipment (rockhopper gear and 40 meter sweep length).
${ }^{2}$ In millions.
Not updated from 2004, new ecosystem survey

Table E2. GREENLAND HALIBUT in Sub-area I and II. Abundance indices from bottom trawl surveys in the Barents Sea and Svalbard area in August (in thousands).

A: The Barents Sea area; B: The expanded Svalbard area.

| A |  | Age |  |  |  |  |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ |  |
| 1995 |  | 42 | - | - | 596 | 989 | 1239 | 1673 | 1020 | - | 195 | - | - | - | 5754 |
| 1996 |  | 12028 | 900 | - | - | - | 415 | 829 | 861 | 85 | 261 | 118 | 82 | - | 15579 |
| 1997 | 1 | 143 | 1162 | 53 | 331 | 589 | 1579 | 2736 | 1120 | 550 | 44 | - | - | - | 8307 |
| 1998 | 1 | 46 | 446 | 328 | 416 | 481 | 323 | 1828 | 924 | 432 | 234 | - | - | - | 5458 |
| 1999 |  | 11637 | 5910 | 384 | 280 | 201 | 1508 | 1729 | 215 | 134 | 661 | 255 | 218 | - | 23132 |
| 2000 |  | - | 619 | 302 | 417 | 816 | 620 | 1163 | 844 | 605 | 270 | 54 | 221 | - | 5931 |
| 2001 |  | - | - | 259 | 203 | 743 | 1120 | 293 | 697 | - | 215 | 107 | - | - | 3637 |
| 2002 |  | - | - | - | 85 | 773 | 2509 | 3047 | 165 | 290 | 839 | - | 255 | - | 7963 |
| 2003 |  | - | - | - | 420 | 450 | 1630 | 1070 | 840 | 250 | 410 | - | - | - | 5070 |
| B |  | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Year |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ |  |
| 1995 |  | 77 | - | - | 429 | 1255 | 1720 | 2535 | 665 | 135 | 281 | 136 | 95 | - | 7328 |
| 1996 |  | 1760 | 360 | 105 | 291 | 1144 | 2717 | 3525 | 1290 | 309 | 603 | 30 | 92 | 45 | 12271 |
| 1997 |  | 593 | 2357 | 311 | 116 | 593 | 3053 | 3019 | 478 | 312 | 20 | - | - | - | 10852 |
| 1998 |  | 2295 | 2836 | 2918 | 540 | 770 | 2477 | 3248 | 1472 | 340 | 346 | 130 | - | 65 | 17437 |
| 1999 |  | 387 | 263 | 1516 | 3095 | 809 | 836 | 2773 | 486 | 333 | 360 | - | 87 | 140 | 11085 |
| 2000 |  | 1976 | 818 | 1280 | 2836 | 3946 | 3216 | 2112 | 1560 | 460 | 199 | - | 95 | - | 18498 |
| 2001 |  | 4659 | 1690 | 1789 | 2517 | 3536 | 2474 | 1889 | 690 | 383 | 773 | 134 | 27 | 50 | 20611 |
| 2002 |  | 2174 | 2475 | 1718 | 2962 | 4291 | 3620 | 4205 | 1031 | 293 | 1267 | 453 | 304 | 212 | 25005 |
| 2003 |  | 1390 | 600 | 1170 | 3510 | 3350 | 4310 | 3470 | 640 | 520 | 150 | 90 | 140 | - | 19340 |

${ }^{1}$ Only Norwegian and international zones covered. Adjusted (according to the mean distribution in the period 1991-1999) to include the Russian EEZ.
Not updated from 2004, new ecosystem survey

Table E3. GREENLAND HALIBUT in Sub-area I and II. Abundance indices on age from the Norwegian stratified bottom trawl survey in August using a hired commercial vessel (numbers in thousands). Trawls were made at $400-1500 \mathrm{~m}$ depth along the continental slope from $68-80^{\circ} \mathrm{N}$.

| Year | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 |  | 2 |  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |  |
| 1994 |  | 0 | 0 | 0 | 1 | 2001 | 16980 | 11008 | 15552 | 6173 | 1241 | 3628 | 1460 | 443 | 129 | 81 | 11 | 58708 |
| 1995 |  | 0 | 0 | 0 | 0 | 1432 | 16945 | 12946 | 20925 | 6737 | 1975 | 4393 | 1385 | 648 | 152 | 103 | 21 | 67662 |
| 1996 |  | 0 | 0 | 0 | 10 | 704 | 13623 | 18538 | 24908 | 8114 | 1473 | 3223 | 820 | 396 | 131 | 100 | 2 | 72042 |
| 1997 |  | 0 | 0 | 0 | 16 | 1446 | 11738 | 17005 | 18927 | 5383 | 1107 | 3261 | 936 | 600 | 87 | 165 | 16 | 60687 |
| 1998 |  | 0 | 0 | 0 | 66 | 1726 | 7868 | 12399 | 23487 | 6243 | 1458 | 4317 | 1238 | 969 | 13 | 183 | 14 | 59981 |
| 1999 |  | 0 | 0 | 0 | 27 | 1300 | 5901 | 15383 | 20209 | 12019 | 1872 | 5913 | 1167 | 1198 | 273 | 183 | 15 | 65460 |
| 2000 |  | 0 | 0 | 0 | 383 | 1920 | 6901 | 10352 | 17885 | 7795 | 5038 | 3284 | 867 | 458 | 204 | 75 | 16 | 55178 |
| 2001 |  | 0 | 10 | 0 | 95 | 986 | 6107 | 15068 | 22584 | 10086 | 3130 | 5442 | 1146 | 1147 | 267 | 180 | 67 | 66315 |
| 2002 |  | 0 | 3 | 3 | 427 | 2492 | 7730 | 10913 | 21660 | 9847 | 6327 | 4248 | 2468 | 1642 | 619 | 208 | 183 | 68767 |
| 2003 |  | 6 | 18 | 8 | 662 | 3972 | 10293 | 14552 | 20438 | 9191 | 4507 | 6388 | 1902 | 1795 | 861 | 253 | 125 | 74963 |
| 2004 |  | 0 | 5 | 5 | 328 | 3637 | 6962 | 12909 | 20674 | 8692 | 3771 | 3908 | 1663 | 2886 | 1276 | 865 | 641 | 68217 |
| 2005 |  | 3 | 24 |  | 2036 | 9170 | 10195 | 13477 | 8785 | 7683 | 4611 | 4388 | 2500 | 2250 | 995 | 401 | 693 | 67210 |

Not updated from 2006 due to new age reading method
Table EA. GREENLAND HALIBUT in Sub-area I and II. Abundance indices on age from the Norwegian bottom trawl survey north and east of Spitsbergen in September (numbers in thousands).

A: Survey area, Russian EEZ excluded B: Including Russian EEZ

| A <br> Year |  | Age |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6+ |  |
| 1996 |  | 15655 | 14510 | 10025 | 3487 | 1593 | 3349 | 48619 |
| 1997 |  | 3415 | 15271 | 14140 | 2803 | 403 | 434 | 36466 |
| 1998 |  | 8482 | 18718 | 9463 | 5161 | 1166 | 932 | 43922 |
| 1999 |  | 5370 | 9074 | 3328 | 2271 | 1492 | 954 | 22489 |
| 2000 |  | 9529 | 16844 | 8007 | 6274 | 1746 | 722 | 43122 |
| 2001 |  | 26206 | 15765 | 4515 | 1767 | 802 | 465 | 49520 |
| 2002 |  | 40186 | 34065 | 15441 | 3862 | 1320 | 556 | 95430 |
| 2003 |  | 49146 | 37344 | 6336 | 3188 | 1035 | 327 | 97376 |
| 2004 | 1 | 15257 | 28540 | 48286 | 12598 | 3562 | 1153 | 109396 |
| 2005 | 1 | 138248 | 23689 | 25989 | 32052 | 6735 | 893 | 227606 |
| $\begin{aligned} & \text { B } \\ & \text { Year } \end{aligned}$ |  | Age |  |  |  |  |  | Total |
|  |  | 1 | 2 | 3 | 4 | 5 | 6+ |  |
| 1998 |  | 10210 | 28020 | 17186 | 6380 | 1551 | 932 | 64279 |
| 1999 |  | 7514 | 16159 | 8045 | 3067 | 2401 | 954 | 38140 |
| 2000 |  | No coverage in Russian EEZ |  |  |  |  |  |  |
| 2001 |  | 38112 | 40377 | 7960 | 4300 | 1215 | 510 | 92475 |
| 2002 |  | 96231 | 58113 | 31500 | 5665 | 1576 | 556 | 193641 |
| 2003 |  | No coverage in Russian EEZ |  |  |  |  |  |  |
| 2004 | 1 | 23560 | 47023 | 77374 | 14081 | 3719 | 1232 | 166989 |
| 2005 | 1 | 253127 | 40975 | 40231 | 40858 | 6955 | 893 | 383039 |

[^9]Table E5. GREENLAND HALIBUT in Sub-area I and II. Abundance indices from three Norwegian bottom trawl surve ys in the Barents Sea in August - September (from 2004 two of them are part of the joint ecosystem survey covering the whole Barents Sea) combined to one index (in thousands).

A: Old strata system used B: Ecosystem survey combined with Norw. GrHal survey

| A | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1996 | 17926 | 14906 | 10134 | 4486 | 16194 | 22217 | 30014 | 10163 | 1857 | 3954 | 957 | 523 | 175 | 100 | 2133608 |
| 1997 | 4050 | 18107 | 14547 | 4481 | 12917 | 20753 | 22984 | 6362 | 1563 | 3312 | 936 | 600 | 87 | 165 | 16110880 |
| 1998 | 10704 | 21705 | 12521 | 7603 | 9915 | 14680 | 27784 | 7800 | 1937 | 4586 | 1353 | 1027 | 13 | 241 | 14121883 |
| 1999 | 5895 | 9451 | 5200 | 7116 | 8412 | 17437 | 24175 | 12857 | 2407 | 6595 | 1294 | 1387 | 273 | 183 | 144102826 |
| 2000 | 11474 | 17755 | 9870 | 11359 | 13093 | 14139 | 20608 | 9704 | 5707 | 3548 | 901 | 695 | 204 | 75 | 16119148 |
| 2001 | 30631 | 17452 | 6521 | 5115 | 10077 | 17548 | 24465 | 10973 | 3440 | 6280 | 1302 | 1147 | 267 | 180 | 67135464 |
| 2002 | 42348 | 36537 | 17472 | 9105 | 13649 | 15040 | 27076 | 10130 | 6679 | 5104 | 2909 | 1893 | 619 | 257 | 183188999 |
| 2003 | 50512 | 37972 | 8298 | 11410 | 15428 | 20553 | 24664 | 10521 | 5437 | 6958 | 1992 | 1955 | 861 | 253 | 125196939 |
| 2004 | 17233 | 29072 | 50471 | 17112 | 13233 | 16459 | 24970 | 9753 | 4568 | 4170 | 1963 | 3042 | 1460 | 865 | 726195096 |
| 2005 | 153834 | 29173 | 32072 | 46345 | 24680 | 20381 | 14189 | 9919 | 5261 | 4929 | 2709 | 2392 | 1242 | 540 | 776348443 |


| B | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 2004 | 16513 | 37564 | 56050 | 12858 | 11967 | 18047 | 25933 | 10060 | 4974 | 4413 | 2151 | 3600 | 1276 | 865 | 641206912 |
| 2005 | 182754 | 40350 | 40139 | 40760 | 25334 | 21739 | 15320 | 10504 | 5594 | 5131 | 2967 | 2494 | 1249 | 686 | 758395780 |

[^10]Table E6. GREENLAND HALIBUT in Sub-area I and II. Russian autumn bottom trawl surveys: Abundance indices at different age (numbers in thousands).

| Year | Age-group |  |  |  |  |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\leq 3$ | 4 | 5 | 56 | $6 \quad 7$ | 78 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |  |
| 1984 | 4124 | 5359 | 7788 | 24951 | 5119863 | 11499 | 6750 | 5416 | 2420 | 1196 | 247 | 146 | 143 | 89902 |
| 1985 | 3331 | 4371 | 17076 | 635648 | 8827826 | 611717 | 5722 | 4090 | 1937 | 895 | 311 | 31 | 131 | 113086 |
| 1986 | 2687 | 6600 | 15853 | 25696 | 616468 | 8436 | 3811 | 2660 | 974 | 539 | 184 | 72 | 6 | 80986 |
| 1987 | 289 | 6761 | 9724 | 12703 | 7633 | 3867 | 1903 | 1627 | 721 | 416 | 110 | 0 | 38 | 45792 |
| 1988 | 2591 | 4409 | 7891 | 14181 | 111311 | 14308 | 2253 | 1756 | 820 | 307 | 125 | 163 | 54 | 5016 c |
| 1989 | 1429 | 11310 | 13124 | 25881 | 112782 | 2989 | 2381 | 1285 | 334 | 271 | 98 | 102 | 118 | 75104 |
| 1990 | 2820 | 8360 | 16252 | 15621 | 111393 | 3120 | 1911 | 1158 | 307 | 198 | 58 | 36 | 0 | 62234 |
| $1991{ }^{1}$ | 1422 | 8455 | 25408 | 21843 | 315235 | 59419 | 2369 | 1211 | 655 | 142 | 95 | 16 | 26 | 86296 |
| 1992 | 685 | 7461 | 33341 | 25498 | 817272 | 210178 | 2720 | 1262 | 938 | 318 | 67 | 0 | 0 | 9974 C |
| 1993 | 114 | 2166 | 13317 | 19752 | 2216528 | 810305 | 3370 | 1868 | 903 | 519 | 103 | 111 | 111 | 69167 |
| 1994 | 49 | 1604 | 9868 | 17549 | 911533 | 7746 | 3401 | 1876 | 605 | 394 | 114 | 114 | 57 | 54 91C |
| 1995 | 19 | 467 | 5759 | 18222 | 2215296 | 611539 | 4393 | 1413 | 529 | 312 | 84 | 11 | 32 | 58076 |
| $1996{ }^{2}$ | 0 | 1670 | 6680 | 18722 | 22774 | 413354 | 8512 | 476 | 284 | 106 | 115 | 36 | 20 | 71 68c |
| 1997 | 235 | 1575 | 4023 | 12165 | 5515919 | 916452 | 4591 | 1432 | 779 | 162 | 271 | 66 | 88 | 57758 |
| 1998 | 3917 | 5542 | 7768 | 15589 | 9 16842 | 17727 | 9676 | 2548 | 1752 | 535 | 254 | 85 | 72 | 82307 |
| 1999 | 4057 | 4961 | 5951 | 12350 | 014255 | 16078 | 7952 | 3009 | 965 | 494 | 307 | 74 | - | 70453 |
| 2000 | 2841 | 5327 | 10718 | 15719 | 918694 | 421235 | 9155 | 3593 | 2580 | 1011 | 108 | 133 | 120 | 91234 |
| 2001 | 1592 | 6884 | 17365 | 37881 | 8127661 | 14163 | 6576 | 3988 | 1875 | 1713 | 929 | 217 | 180 | 121024 |
| $2002^{3}$ | 2145 | 7127 | 10771 | 44220 | 033675 | 518747 | 5947 | 5477 | 1216 | 1877 | 1973 | 60 | 120 | 133355 |
| 2003 | 1735 | 6479 | 10029 | 19751 | 114160 | 7592 | 3519 | 2555 | 2200 | 1664 | 831 | 141 | 470 | 71126 |
| 2004 | 3305 | 8342 | 9461 | 21834 | 8422876 | 614187 | 8331 | 3776 | 2544 | 1745 | 1031 | 811 | 966 | 99 20c |
| 2005 | 2096 | 7668 | 11657 | 17933 | 33555 | 514140 | 4658 | 3264 | 1844 | 1585 | 789 | 554 | 420 | 87164 |
| 2006 | 3099 | 13954 | 18873 | 34869 | 9937481 | 120542 | 7631 | 3586 | 2489 | 2329 | 1663 | 720 | 785 | 148021 |
| 2007 | 995 | 5713 | 15982 | 27722 | 2236544 | 18917 | 9382 | 6033 | 5221 | 5171 | 2297 | 1399 | 1134 | 13651 C |
| 2008 | 1483 | 11642 | 12475 | 21157 | 572551 | 133844 | 19618 | 6297 | 7262 | 6994 | 5474 | 3240 | 4092 | 166129 |
| 200c | 713 | 13726 | 35041 | 43719 | 940611 | 138274 | 13509 | 4006 | 7371 | 4522 | 4152 | 1257 | 1398 | 208300 |

${ }^{1}$ Age composition based on combined age-length-keys for 1990 and 1992.
${ }^{2}$ Only half of standard area investigated.
${ }^{3}$ Adjusted assuming area distibution as in 2001.

Table E7. GREENLAND HALIBUT catch in weight, numbers, and biomass (in tonnes) and abundance (in thousands) estimated from Spanish survey 1997-2008.

| Year | Catch $(\mathrm{Kg})$ | Catch (numbers) | Biomass $^{\mathrm{TM}}$ | Abundance ('000) |
| :---: | :---: | :---: | :---: | :---: |
| 1997 | 195056 | 211533 | 344014 | 379444 |
| 1998 | 180974 | 187259 | 351466 | 373149 |
| 1999 | 198781 | 172687 | 436956 | 377792 |
| 2000 | 169389 | 140355 | 340619 | 291265 |
| 2001 | 152681 | 129289 | 283511 | 249219 |
| 2002 | 144335 | 115213 | 256460 | 207466 |
| 2003 | 151952 | 132117 | 283644 | 256327 |
| 2004 | 153859 | 135631 | 320485 | 283965 |
| 2005 | 144573 | 134566 | 317320 | 313459 |
| $2006^{*}$ |  |  |  |  |
| $2007^{*}$ |  | 101578 | $379456 / 129221^{* *}$ | $424822 / 144561^{* *}$ |
| 2008 | 91573 |  |  |  |

[^11]Table E8. GREENLAND HALIBUT in Sub-area I and II. Abundance indices from bottom trawl surveys in the Barents Sea in winter (in thousands).

A: Restricted area surveyed every year; B: Enlarged area (includes the restricted one) surveyed since 1993

| A | Year | Age |  |  |  |  |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ |  |
|  | 1989 | 1078 | 788 | 1056 | 2284 | 3655 | 2655 | 864 | 971 | 210 | - | 19 | 76 | 56 | 13712 |
|  | 1990 | 66 | 907 | 2071 | 1716 | 1996 | 2262 | 1046 | 365 | 175 | - | 30 | 119 | 165 | 10918 |
|  | 1991 | - | 279 | 755 | 1323 | 1257 | 1526 | 2440 | 906 | 450 | 457 | - | 55 | 127 | 9575 |
|  | 1992 | 63 | 128 | 719 | 897 | 1554 | 543 | 1069 | 791 | - | 648 | 135 | 40 | 53 | 6640 |
|  | 1993 | - | 17 | 168 | 502 | 1730 | 868 | 1490 | 758 | 88 | 655 | 382 | 31 | 35 | 6724 |
|  | 1994 | - | 16 | 142 | 1178 | 2259 | 1644 | 1750 | 885 | - | 506 | 38 | 25 | - | 8443 |
|  | 1995 | - | - | - | 168 | 786 | 749 | 1331 | 760 | 359 | 486 | 60 | 199 | - | 4898 |
|  | 1996 | 1816 | - | 28 | 40 | 709 | 1510 | 2964 | 1000 | 307 | 808 | 154 | 152 | 45 | 9533 |
|  | 1997 | - | 21 | - | 21 | 176 | 812 | 1788 | 1440 | 653 | 209 | 94 | 73 | - | 5287 |
|  | 1998 | - | - | - | 67 | 474 | 1172 | 2491 | 1144 | 302 | 401 | 89 | 19 | 4 | 6163 |
|  | 1999 | - | 77 | 276 | 243 | 495 | 485 | 1058 | 555 | 408 | 152 | 75 | 56 | - | 3880 |
|  | 2000 | - | 40 | 56 | 396 | 719 | 519 | 1187 | 261 | 290 | 531 | 131 | 23 | 55 | 4208 |
|  | 2001 | 19 | 36 | 112 | 558 | 517 | 260 | 497 | 697 | 267 | 478 | 43 | 42 | 30 | 3556 |
|  | 2002 | - | - | 32 | 609 | 1019 | 1148 | 989 | 362 | 139 | 591 | 106 | 54 | 54 | 5103 |


| B | Year | Age |  |  |  |  |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13+ |  |
|  | 1993 | - | 17 | 279 | 1002 | 3129 | 2818 | 3895 | 1632 | 309 | 1406 | 616 | 31 | 35 | 15169 |
|  | 1994 | - | 16 | 152 | 1482 | 3768 | 2698 | 3420 | 1615 | - | 1171 | 135 | 25 | - | 14482 |
|  | 1995 | - | - | - | 216 | 2824 | 6229 | 10624 | 2727 | 1250 | 1902 | 172 | 718 | 57 | 26719 |
|  | 1996 | 3149 | - | 28 | 102 | 1547 | 3043 | 4991 | 1599 | 472 | 1211 | 317 | 250 | 72 | 16781 |
|  | $1997{ }^{1}$ | - | 163 | - | 203 | 624 | 2742 | 5759 | 4170 | 1653 | 562 | 240 | 181 | 66 | 16363 |
|  | $1998{ }^{1}$ | 220 | 501 | 2797 | 1011 | 1847 | 3477 | 6539 | 3057 | 867 | 1179 | 301 | 96 | 57 | 21949 |
|  | 1999 | 41 | 195 | 691 | 825 | 829 | 1531 | 3130 | 1496 | 1011 | 500 | 115 | 129 | 101 | 10594 |
|  | 2000 | 169 | 482 | 947 | 5425 | 2575 | 1310 | 3035 | 553 | 796 | 1109 | 284 | 27 | 55 | 16767 |
|  | 2001 | 69 | 250 | 363 | 2046 | 4250 | 2730 | 2983 | 1123 | 416 | 1148 | 111 | 137 | 94 | 15720 |
|  | 2002 | 233 | 104 | 248 | 1373 | 2748 | 3265 | 3641 | 932 | 449 | 1714 | 365 | 177 | 178 | 15427 |
|  | 2003 | 50 | 89 | 151 | 785 | 1786 | 2860 | 5411 | 1313 | 289 | 951 | 356 | 189 | 92 | 14322 |
|  | 2004 | 67 | 118 | 128 | 527 | 1294 | 1099 | 3207 | 1220 | 624 | 504 | 201 | 281 | 266 | 9536 |
|  | 2005 | 259 | 300 | 2318 | 1512 | 4106 | 3554 | 5373 | 2072 | 862 | 278 | 372 | 305 | 824 | 22135 |
|  | 2006 | 45 | 46 | 1119 | 5518 | 6912 | 5640 | 1353 | 603 | 562 | 321 | 365 | 61 | 115 | 22660 |

${ }^{1}$ Adjusted (according to the 1996 distribution) to include the Russian EEZ which was not covered by the survey.
Not updated from 2007 due to new age reading method

Table E9 GREENLAND HALIBUT in Sub-areas I and II. Results from a research program using trawlers in a limited commercial fishery 1992-2005. All areas combined. Spring and autumn combined in 1992-1993, otherwise only spring-data.

| Catch in numbers on age (\%) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :---: |
| Age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 3 | 0.1 |  |  | 0.1 |  | 0.0 | 0.0 | 0.0 |  |  |  |  | 0.1 | 0.2 |  |
| 4 | 4.6 | 4.2 | 3.2 | 0.7 | 0.5 | 0.9 | 0.2 | 0.7 | 1.2 | 1.3 | 0.7 | 1.8 | 1.4 | 1.8 |  |
| 5 | 19.1 | 25.0 | 24.7 | 22.5 | 19.5 | 24.8 | 6.6 | 7.7 | 10.8 | 6.3 | 7.7 | 8.5 | 8.9 | 5.4 |  |
| 6 | 23.0 | 18.4 | 23.8 | 22.6 | 31.6 | 22.9 | 25.5 | 23.0 | 17.1 | 20.2 | 16.8 | 21.7 | 18.9 | 20.4 |  |
| 7 | 25.9 | 27.1 | 26.8 | 30.2 | 35.6 | 30.5 | 44.5 | 39.6 | 43.0 | 28.5 | 42.5 | 30.5 | 31.3 | 25.4 |  |
| 8 | 13.3 | 12.4 | 11.2 | 11.0 | 8.7 | 10.1 | 15.5 | 14.5 | 12.3 | 24.5 | 12.4 | 9.6 | 14.8 | 21.5 |  |
| 9 | 1.7 | 0.7 | 1.0 | 2.7 | 1.3 | 2.6 | 4.5 | 1.6 | 4.5 | 7.8 | 7.1 | 8.1 | 9.5 | 8.2 |  |
| 10 | 6.8 | 7.4 | 5.9 | 6.6 | 2.0 | 5.0 | 2.0 | 9.7 | 8.5 | 7.3 | 8.8 | 11.0 | 4.7 | 6.5 |  |
| 11 | 2.9 | 3.1 | 2.4 | 2.0 | 0.5 | 1.9 | 0.8 | 1.0 | 0.9 | 1.9 | 2.2 | 4.1 | 4.0 | 3.1 |  |
| 12 | 1.7 | 1.0 | 0.6 | 1.1 | 0.2 | 0.8 | 0.3 | 1.8 | 1.1 | 1.7 | 1.2 | 3.1 | 3.5 | 4.0 |  |
| 13 | 0.5 | 0.4 | 0.2 | 0.3 | 0.0 | 0.3 |  | 0.2 | 0.6 | 0.3 | 0.2 | 1.2 | 1.5 | 2.1 |  |
| 14 | 0.2 | 0.2 | 0.1 | 0.2 | 0.1 | 0.2 |  | 0.2 | 0.0 | 0.2 | 0.4 | 0.5 | 0.9 | 1.0 |  |
| 15 | 0.1 |  |  |  |  | 0.0 |  | 0.0 | 0.0 | 0.2 | 0.1 | 0.0 | 0.4 | 0.5 |  |


| Mean individual weight (kg) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 3 | 0.26 |  |  | 0.40 |  | 0.39 |  |  |  |  |  |  | 0.27 | 0.24 |
| 4 | 0.50 | 0.53 | 0.52 | 0.47 | 0.48 | 0.45 | 0.41 | 0.51 | 0.50 | 0.60 | 0.44 | 0.48 | 0.44 | 0.48 |
| 5 | 0.71 | 0.76 | 0.73 | 0.70 | 0.74 | 0.69 | 0.76 | 0.74 | 0.69 | 0.66 | 0.69 | 0.68 | 0.65 | 0.64 |
| 6 | 0.96 | 0.98 | 0.95 | 0.94 | 0.94 | 0.88 | 0.96 | 0.92 | 0.98 | 0.94 | 0.93 | 1.00 | 0.88 | 0.84 |
| 7 | 1.29 | 1.33 | 1.28 | 1.24 | 1.23 | 1.15 | 1.19 | 1.25 | 1.23 | 1.12 | 1.22 | 1.28 | 1.17 | 1.14 |
| 8 | 1.77 | 1.85 | 1.79 | 1.71 | 1.66 | 1.55 | 1.79 | 1.64 | 1.57 | 1.48 | 1.39 | 1.67 | 1.43 | 1.40 |
| 9 | 2.00 | 2.28 | 2.23 | 2.03 | 2.00 | 1.87 | 2.26 | 2.18 | 1.90 | 1.84 | 1.69 | 1.97 | 1.73 | 1.67 |
| 10 | 2.46 | 2.65 | 2.55 | 2.50 | 2.50 | 2.34 | 2.54 | 2.38 | 2.40 | 2.30 | 2.31 | 2.37 | 2.14 | 2.26 |
| 11 | 3.10 | 3.43 | 3.37 | 3.28 | 3.16 | 2.95 | 3.47 | 3.17 | 3.13 | 2.92 | 3.19 | 3.20 | 2.34 | 2.62 |
| 12 | 3.86 | 4.32 | 4.22 | 3.71 | 3.70 | 3.46 | 4.16 | 3.79 | 4.04 | 3.82 | 3.91 | 3.48 | 2.77 | 2.87 |
| 13 | 4.44 | 5.18 | 5.01 | 4.62 |  | 4.52 |  | 5.07 | 4.47 | 3.68 | 5.20 | 4.28 | 2.92 | 2.98 |
| 14 | 6.00 | 6.44 | 6.29 | 5.59 |  | 5.47 |  | 5.60 | 6.00 | 5.74 | 5.59 | 4.74 | 3.89 | 3.30 |
| 15 | 5.22 |  |  |  |  |  |  |  | 8.79 | 5.52 | 7.03 | 9.17 | 4.65 | 3.32 |

Not updated from 2006 due to new age reading method

Table E9 (Continued) GREENLAND HALIBUT in Sub-areas I and II. Results from a research program using trawlers in a limited commercial fishery 1992-2005. All areas combined. Spring and autumn combined in 1992-1993, otherwise only spring-data.

| CPUE (N) on age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 3 | 0 |  |  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 |  |
| 4 | 19 | 30 | 26 | 7 | 7 | 11 | 2 | 7 | 14 | 12 | 7 | 19 | 15 | 24 |
| 5 | 80 | 176 | 198 | 219 | 286 | 298 | 59 | 72 | 132 | 63 | 81 | 90 | 96 | 70 |
| 6 | 97 | 130 | 191 | 220 | 463 | 275 | 229 | 214 | 208 | 201 | 176 | 229 | 203 | 263 |
| 7 | 109 | 191 | 215 | 294 | 521 | 366 | 400 | 369 | 524 | 284 | 447 | 322 | 337 | 328 |
| 8 | 56 | 87 | 90 | 107 | 127 | 121 | 139 | 135 | 150 | 244 | 130 | 101 | 159 | 278 |
| 9 | 7 | 5 | 8 | 26 | 19 | 31 | 40 | 15 | 55 | 78 | 75 | 86 | 102 | 106 |
| 10 | 29 | 52 | 47 | 64 | 29 | 60 | 18 | 90 | 104 | 73 | 92 | 116 | 51 | 84 |
| 11 | 12 | 22 | 19 | 19 | 7 | 23 | 7 | 9 | 11 | 18 | 23 | 43 | 43 | 40 |
| 12 | 7 | 7 | 5 | 11 | 3 | 10 | 3 | 17 | 13 | 17 | 12 | 32 | 38 | 52 |
| 13 | 2 | 3 | 2 | 3 | 0 | 4 | 0 | 2 | 7 | 3 | 2 | 12 | 16 | 27 |
| 14 | 1 | 1 | 1 | 2 | 1 | 2 | 0 | 2 | 0 | 2 | 4 | 5 | 10 | 13 |
| 15 | 0 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 4 | 6 |


| CPUE (kg) on age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |  |
| 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 3 | 0 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |  |  |  |  |
| 4 | 10 | 16 | 13 | 3 | 4 | 5 | 1 | 3 | 7 | 7 | 3 | 9 | 6 | 11 |  |
| 5 | 57 | 134 | 145 | 153 | 211 | 207 | 45 | 53 | 91 | 41 | 56 | 61 | 63 | 44 |  |
| 6 | 93 | 127 | 182 | 207 | 435 | 243 | 220 | 197 | 204 | 189 | 164 | 229 | 179 | 220 |  |
| 7 | 140 | 254 | 276 | 364 | 641 | 423 | 476 | 461 | 645 | 318 | 543 | 411 | 396 | 373 |  |
| 8 | 99 | 162 | 161 | 183 | 211 | 189 | 249 | 221 | 236 | 361 | 181 | 169 | 228 | 389 |  |
| 9 | 14 | 11 | 18 | 53 | 38 | 59 | 91 | 32 | 105 | 143 | 127 | 169 | 177 | 176 |  |
| 10 | 70 | 138 | 121 | 161 | 73 | 141 | 46 | 215 | 250 | 167 | 213 | 275 | 109 | 189 |  |
| 11 | 38 | 75 | 65 | 64 | 23 | 68 | 25 | 30 | 33 | 54 | 74 | 138 | 101 | 104 |  |
| 12 | 28 | 30 | 20 | 40 | 11 | 33 | 11 | 64 | 53 | 66 | 48 | 113 | 105 | 150 |  |
| 13 | 9 | 15 | 8 | 13 | 0 | 16 | 0 | 9 | 32 | 11 | 9 | 52 | 48 | 79 |  |
| 14 | 5 | 9 | 5 | 11 | 0 | 13 |  | 10 | 2 | 10 | 24 | 23 | 38 | 43 |  |
| 15 | 2 |  |  | 0 | 0 | 0 |  | 0 | 3 | 11 | 4 | 4 | 20 | 20 |  |


|  | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Overall mean individual weight (kg) | 1.35 | 1.38 | 1.27 | 1.29 | 1.12 | 1.16 | 1.30 | 1.39 | 1.35 | 1.38 | 1.38 | 1.57 | 1.37 | 1.39 |
| CPUE (kg round weight per trawlhour)** | 567 | 973 | 1020 | 1255 | 1640 | 1393 | 1169 | 1294 | 1647 | 1377 | 1449 | 1657 | 1475 | 1795 |
| CPUE (Number fish per trawlhour)** | 420 | 705 | 803 | 973 | 1464 | 1201 | 899 | 931 | 1220 | 998 | 1050 | 1055 | 1077 | 1291 |
| Catch (in tonnes) | 695 | 862 | 811 | 368 | 436 | 274 | 272 | 269 | 295 | 297 | 288 | 298 | 304 | 292 |

*) Preliminary

* *) Average for freezer- and factorytrawler

Not updated from 2006 due to new age reading method

Table E10. GREENLAND HALIBUT in ICES Sub-area IV (North Sea. Nominal catch (t) by countries as officially reported to ICES. Not included in the assessment.

| Year | Denmark | Faroe <br> Islands | France | Germany | Green- <br> land | Ire- <br> land | Norway | Russia | UK <br> England <br> \& Wales | UK <br> Scotland | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1973 | - | - - | - | 4 |  | - - | 9 |  | $8 \quad 28$ | - | 49 |
| 1974 | - | - | - | 2 |  | - - | 2 |  | - 30 | - | 34 |
| 1975 | - | - - | - | 1 |  | - - | 4 |  | - 12 | - | 17 |
| 1976 | - | - - | - | 1 |  | - - | 2 |  | 18 | - | 21 |
| 1977 | - | - - | - | 2 |  | - - | 2 |  | - 8 | - | 12 |
| 1978 | - | - - | 2 | 30 |  | - . | - |  | 1 | - | 33 |
| 1979 | - | - - | 2 | 16 |  | - - | 2 |  | - 1 | - | 21 |
| 1980 | - | - 177 | - | 34 |  | - . | 5 |  | - - | - | 216 |
| 1981 | - | - - | - | - |  | - - | 7 |  | - - | - | 7 |
| 1982 | - | - - | 2 | 26 |  | - - | 17 |  | - - | - | 45 |
| 1983 | - | - - | 1 | 64 |  | - - | 89 |  | - | - | 154 |
| 1984 | - | - - | 3 | 50 |  | - - | 32 |  | - - | - | 85 |
| 1985 | - | 1 | 2 | 49 |  | - - | 12 |  | - - | - | 64 |
| 1986 | - | - | 30 | 2 |  | - - | 34 |  | - | - | 66 |
| 1987 | - | 28 | 16 | 1 |  | - - | 35 |  | - | - | 80 |
| 1988 | - | 71 | 62 | 3 |  | - - | 19 |  | 1 | - | 156 |
| 1989 | - | 21 | $14^{1}$ | 1 |  | - - | 197 |  | - 5 | - | 238 |
| 1990 | - | 10 | $30^{1}$ | 3 |  | - - | 29 |  | - 4 | - | 76 |
| 1991 | - | 48 | $291{ }^{1}$ | 1 |  | - . | 216 |  | - 2 | - | 558 |
| 1992 | 1 | 15 | $416{ }^{1}$ | 3 |  | - - | 626 |  | - + | 1 | 1062 |
| 1993 | 1 | - | $78^{1}$ | 1 |  | - - | 858 |  | 10 | + | 948 |
| 1994 | + | - 103 | $84^{1}$ | 4 |  | - - | 724 |  | - 6 | - | 921 |
| 1995 | + | 706 | 165 | 2 |  | - - | 460 |  | - 52 | 283 | 1668 |
| 1996 | + | + | 249 | 1 |  | - . | 1496 |  | - 105 | 159 | 2010 |
| 1997 | + | + | 316 | 3 |  | - - | 873 |  | - 1 | 162 | 1355 |
| 1998 | + | - | $71^{1}$ | 10 |  | 10 | 804 |  | - 35 | 435 | 1365 |
| 1999 | + | - |  | 1 |  | 18 | 2157 |  | - 43 | 358 | 2577 |
| 2000 | + | + | 41 | 10 |  | 19 | $498{ }^{1}$ |  | - 67 | 192 | 827 |
| 2001 | + | + | 43 | - |  | 10 | 470 |  | - 122 | 202 | 847 |
| 2002 | + | + | 8 | + |  | 2 | 200 |  | - 10 | 246 | 466 |
| 2003 | - | - - | 1 | + |  | + | 453 |  | - + | 122 | 576 |
| 2004 | - | - - | - | - |  | - . | 413 |  | - 90 | - | 503 |
| 2005 | - | - - | 2 | - |  | - - | 58 |  | - 4 | - | 64 |
| 2006 | - | - - | 3 | - |  | - | 89 |  | - 7 | - | 99 |
| 2007 | 1 - | + | + | - |  | - - | 129 |  | - + | + | 129 |
| 2008 | 1 - | - - | - | - |  | - - | 14 |  | - 22 | - | 36 |
| 2009 | 1 - | - - | - | - |  | - | 5 |  | - 129 | - | 134 |

[^12]

Figure 8.1. NEA Greenland halibut. Log catchability residuals by age and year for the tuning fleets included in the assessments. For each graph all bubbles are normalized to the same maximum bubble-size. Open bubbles represent positive values; filled bubbles represent negative values.


Figure 8.2. NEA Greenland halibut. Historical landings, recruitment, fishing mortality and spawning stock biomass.


Figure 8.3. NEA Greenland halibut. Retrospective plots.


Figure 8.4. NEA Greenland halibut. Biomass estimates from the tuning series used in the assessment. Years with open symbols in the Russian series excluded from the tuning. The Norwegian CP UE Survey was ended in 2006.


Figure 8.5. NEA Greenland halibut. Swept area estimate of the mature female biomass based on the data from the Norwegian Greenland halibut surve $y$ along the continental slope (August) and Russian trawl survey (October-December).


Figure 8.6 XSA estimates for recruitment, spawning stock biomass (SSB) and fisheries mortality (F) for NEA-Greenland halibut as found in the latest ICES assessment (stippled black line) (ICES 2009 AFWG report), with plus group changed from age 15 to 13 (grey diamonds), catchability constant in the analysis from age 11 in stead of 10 (grey open circles), and a run with both 13 as plus group and catbchability constant in the analysis from age 11 (black solid line) (see Hallfredsson 2010).

### 9.1 Regulation of the Barents Sea Capelin Fishery

Since 1979, the Barents Sea capelin fishery has been regulated by a bilateral fishery management agreement between Russia (former USSR) and Norway. A TAC has been set separately for the winter fishery and for the autumn fishery. In recent years (from 1999) no autumn fishery has taken place, except for a small Russian experimental fishery. The fishery was closed from 1 May to 15 August until 1984. After 1984, the fishery was closed from 1 May to 1 September. A minimum landing size of 11 cm has been in force for years of regulating fishery. From the autumn of 1986 to the winter of 1991, from the autumn 1993 to the $w$ inter 1999, and in 2004-2008, no commercial fishery took place. A commercial fishery in the wintering-spring period started again in 2009. AFWG strongly recommends capelin fishery only on mature fish and during the period from January to April only.

### 9.2 Catch Statistics (Table 9.1, 9.2)

The total catches that were taken during spring 2010 amounted to 246209 tonnes to Norway and 77367 tonnes to Russia. Russian catch statistics are shown in Tables 9.1. The age-length composition showed some variation in time and place of fishery. Because of this, five regions for length-age calculation of catch statistic for Russian fleet were used.
Data of age-length composition is presented in table 9.1. The international historical catch by country and seasons in the years 1972-2009 is given in Table 9.2.

### 9.3 Sampling

The sampling from scientific surveys, exploratory fishing and observers of capelin from January 2009 - April 2010 is summarised below:

| Investigation | No. of samples | Length <br> measurements | Aged <br> individuals |
| :--- | :--- | :--- | :--- |
| Capelin investig ations winter 2009 (Norway) | 103 | 26805 | 3137 |
| Capelin investig ations winter 2009 (Russia) | 46 | 5529 | 710 |
| Exploratory fishing winter 2009 (Russia) | 101 | 28958 | 700 |
| Bottom survey winter 2009 (Russia) | 26 | 1511 | - |
| Bottom survey winter 2009 (Norway) | 193 | 6125 | 2625 |
| Young herring surv. In the Barents Sea, May 2009 <br> (Russia) | 14 | 2414 | 484 |
| Ecosystem survey autumn 2009 (Norway) | 308 | 16953 | 3546 |
| Ecosystem survey autumn 2009 (Russia) | 458 | 13868 | 1142 |
| Bottom fish survey, November 2009 (Russia) | 144 | 10799 | 275 |
| Exploratory polar cod fishing autumn 2009 (Russia) | 27 | 3974 | 200 |
| Capelin winter investig ations 2010 (Russia) | 27 | 9598 | 1100 |
| Observer on fishing vessels in winter-spring | 90 | 11878 | 1000 |
| 2010(Russia) |  |  |  |
| Bottom survey winter 2010 (Norway) | 133 | 8950 | 3810 |
| Bottom survey winter 2010 (Russia) | 46 | 3151 | 50 |
| Sampling from fishing vessels in winter-spring 2010   <br> (Norway) in processing in processing | in <br> Total 2009- 2010 | 1716 | 150513 |

### 9.4 Stock Size Estimates

### 9.4.1 Acoustic stock size estimates in 2009 (Table 9.3)

One Russian and three Norwegian vessels jointly carried out the 2009 acoustic survey as part of an ecosystem survey during autumn (Anon., 2009). The geographical coverage of the total stock was considered complete. It was synoptic as in the previous year and the results of estimation are representative. The geographical distribution of capelin is shown in Figure 9.1.
The total capelin stock was estimated at 3.76 million tonnes. It is about $15 \%$ lower than the stock estimated last year but higher than the long term mean level. Almost $62 \%$ ( 2.3 million tonnes) of the stock biomass consisted of maturing fish ( $>14.0 \mathrm{~cm}$ ). The estimated maturing stock is some smaller then in 2008. The results from the survey are given in Table 9.3.

### 9.4.2 Recruitment estimation in 2009 (Table 9.4)

The historical estimated total number of larvae is shown in Table 9.4. These larval abundance estimates should reflect the amount of larvae produced each year (Gundersen and Gjøsæter, 1998). There were some problems with this survey in 1986, 1995 and since 1997 when permission has not been granted to enter the Russian EEZ . During the last three years the larval surveys based on Gulf III plankton samples, which have been carried out in June each year since 1981, were not conducted.

A swept volume index (Dingsør, 2005; Eriksen et al., 2007) of abundance of 0-group capelin in August-September is given in Table 9.4. This index is calculated both without correction and with correction for catching efficiency (Anon. 2007). The four successive and abundant years classes of capelin were last years and 2009 year class is rich.

Table 9.4 also shows the number of fish in the various year classes, and their "survey mortality" from age one to age two. As there has been no fishing on these age groups, the figures for total mortality constitute natural mortality only, and probably reflect quite well the predation on capelin.

### 9.5 Other surveys and information from 2009-2010

### 9.5.1 Russian capelin spring investigation

Data on capelin prespawning concentrations in the wintering grounds, the pattern of prespawning migrations, periods and areas of fish approaches to the coasts for spawning were obtained using the results of fishing vessel activity as well as the data from the cruise by R/V "Vilnyus" (02.02-03.03) and from the scientific observers onboard fishing vessels "Admiral Shabalin" (26.01-17.03), "Capitan Morgun" (04-24.02) and "Demjansk" (27.03-06.04 and 08-10.04 ).

More details and information about fishery and spring investigation is possible to find in WD02.

### 9.5.2 Norwegian capelin spring investigation.

No special capelin investigation was conducted by Norway in winter-spring 2010. The threeyear program to investigate the possibilities for implementing stock size estimates obtained during winter in the management of capelin, was ended in 2009,
and was reported on in 2010. The conclusion was that it is not advisable to base the quotas on stock size estimates from the winter period, since such estimates are much more uncertain than those obtained during autumn.

The biological samples used for calculating the Norwegian catch in numbers by age are collected by 6 purse seiners taking part in the commercial catch. These vessels belong to the Norw egian Reference Fleet.

### 9.6 Stock development assessment

As decided by the Arctic Fisheries Working Group at its 2009 meeting (ICES 2009), the assessment of Barents Sea capelin was left to the parties responsible for the autumn survey, i.e. IMR in Bergen and PINRO in Murmansk. In accordance with this, the assessment was made during a meeting in Kirkenes after the survey. The assessment was an update assessment, without changes in the methodology.

Estimates of stock in number by age group and total biomass for the historical period are shown in Table 9.5. Other data which were used for stock development assessment are shown in table 9.6.

A probabilistic projection of the spawning stock to the time of spawning at 1 April 2010 was made using the spreadsheet model CapTool (implemented in the @RISK add-on for EXCEL, 15000 simulations were used). The projection was based on a maturation and predation model with parameters estimated by the model Bifrost and data on cod abundance and size at age from the 2009 Arctic Fisheries Working Group. The methodology is described in "Stock assessment methodology for the Barents Sea capelin", WD22, AFWG 2008. The predation model for the period JanuaryMarch was based on data from the period 1983-2002. It was decided to draw the natural mortality during October-December randomly from estimates for the period 1995-2001. Also, drawing from the entire period 1983-2002 would include some years with very high estimated natural mortality based on low stock sizes. The models for maturation, predation and mortality are unchanged since 2003.

Probabilistic prognoses for the maturing stock from October 12009 until April 12010 were made, with a CV of 0.20 on the abundance estimate. A CV of 0.20 is slightly higher than the value calculated for most years in Tjelmeland (WD1, 2008). With no catch, the estimated median spawning stock size in 2010 is 750000 tonnes. With a catch of 360000 tonnes, the probability for the spawning stock in 2009 to be below 200 000 t , the $\mathbf{B}_{\mathrm{lim}}$ value used by ACFM in recent years, is $5 \%$ (Fig. 9.2). The median spawning stock size in 2010 will then be 504000 tonnes. Fig 9.2 shows the $95 \%$ percentile of the spawning stock biomass 1 April 2010 as a function of the quota, while Fig 9.3 shows the probability of $\mathrm{SSB}<\mathrm{B}_{\lim }$ as a function of the catch. The monthly distribution of the catch was assumed to be 20 \% in January, 30 \% in February and 50 \% in March. A 1.5-year prognosis has been made for this stock in 2008. Such a prognosis was not carried out this year. Instead, we will give a qualitative view on how the stock will develop in the coming years. This view is to a large extent based on the observation that the three capelin stock collapses observed during the last 30 years have all been preceded by a period of high herring abundance in the Barents Sea. However, some years with good capelin recruitment despite high young herring abundance have also been observed (Fig. 9.5).

The 0-group index for herring in 2009 is low, and the ecosystem survey in 2009 also showed that the abundance of age 1-2 herring in the Barents Sea is low (Anon.2009). The total abundance of 1 year and older herring in the Barents Sea in 2009 will thus
be low, and the recruitment conditions for capelin can then be expected to be good in 2010.

The 2009 year class was found to be strong at the 0 -group stage; more than twice the average index. However, the 2008 year class, that had an 0 -group index more than three times the average value, was reduced to about $75 \%$ of average numbers at age 1 (Table. 9.4). Uncertainties in the survey estimates might partly explain this large reduction, but if this large reduction represents a real increased mortality, this may cause a reduction in the total stock in the years to come.
Being a forage fish in an ecosystem where two of its predators cod and haddock are presently at historic high levels, the capelin stock is now under heavy predation pressure. Consumption estimates from recent years indicate that the amount of capelin consumed by cod (table 1.3,1.4) and haddock (Dolgov, WD\#04) has increased and is at historic high levels. At the same time, capelin during the two last years reached levels where, the current harvest control rule allowed a capelin fishery to take place (table 9.5). Consequently, the stock is under "double pressure" and should be monitored carefully to look for signs of overexploitation that could, eventually, lead to recruitment failure and a reduced stock size. The fishing operations should also be monitored carefully to check whether additional mortality caused by slipping, sorting through the meshes etc. could be a potential problem

### 9.7 Reference points

A Blim (SSBlim) management approach has been suggested for this stock (Gjøsæter et al. 2002). In 2002, the Mixed Russian-Norwegian Fishery Commission agreed to adopt a management strategy based on the rule that, with $95 \%$ probability, at least 200000 t of capelin should be allowed to spawn. Consequently, 200000 t was used as a Blim . There is clearly also a need for a target biomass reference point for capelin, and calculations of Btarget are also in progress.

### 9.8 Regulation of the fishery for 2010

During its autumn 2009 meeting, the Joint Russian-Norwegian Fishery Commission decided that the quota according to the harvest control rule in 2010 will be 360000 tonnes, of which 10000 tonnes ( 5000 tonnes to Norway and 5000 tonnes to Russia) is a research quota.

### 9.9 The Barents Sea capelin benchmark assess ment 2009

In August 2009 a benchmark assessment workshop for shortlived species (WKSHORT) was arranged in Bergen, Norway, and the Barents Sea capelin stock was among the stocks dealt with during that workshop (ICES 2009a). In the report it is stated:

The data and methodology used for the Barents Sea capelin assessment is endorsed by the WKSHORT, based on the combination of available background materials, presentations, discussions, and the draft Report and Stock Annex. Unfortunately, the WKSHORT cannot formally endorse the written version of the approach which appears in the WKSHORT Report and the Stock Annex as of the completion of the WKSHORT on September 4, 2009, as it is incomplete. The WKSHORT is confident that if the Report and Stock Annex canfully convey in writing the information provided throughout the WKSHORT, the Report and Stock Annex will be acceptable.

The WKSHORT endorses the way in which the Barents Sea capelin assessment has incorporated predator-prey interactions (specifically having identified the crucial role of cod predation on capelin mortality rate), and we would suggest that this is world-leading in development of an ecosystem approach. Similarly, the incorporation of uncertainty (through bootstrapping simulations) is to be applauded and has clearly been very effective.
Work is now going on to finalize the Stock Annex, and it will be added to the assessment report made during autumn 2010. It will also be included in the 2011 AFWG report.

Table 9.1 Barents Sea Capelin. Russian catch statistic table. Catch in number ( $10^{6} \mathrm{sp}$.) and biomass (tonnes) by age and length during the fishery in January-April 2010.

| Length, cm | Age/year class |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1(2009) |  | 2(2008) |  | 3(2007) |  | 4(2006) |  | 5(2005) |  | Total(2009-2005) |  |  |  |
|  | N | B | N | B | N | B | N | B | N | B | N | B | N(\%) | B (\%) |
| 5.0 | 2.34 | 0.68 |  |  |  |  |  |  |  |  | 2.34 | 0.68 | 0.07 | + |
| 5.5 | 6.46 | 2.63 |  |  |  |  |  |  |  |  | 6.46 | 2.63 | 0.18 | + |
| 6.0 | 9.01 | 4.94 |  |  |  |  |  |  |  |  | 9.01 | 4.94 | 0.25 | 0.01 |
| 6.5 | 17.19 | 12.85 |  |  |  |  |  |  |  |  | 17.19 | 12.85 | 0.48 | 0.02 |
| 7.0 | 5.21 | 4.92 | 2.10 | 1.98 |  |  |  |  |  |  | 7.31 | 6.90 | 0.21 | 0.01 |
| 7.5 | 2.41 | 2.88 | 2.41 | 2.88 |  |  |  |  |  |  | 4.83 | 5.76 | 0.14 | 0.01 |
| 8.0 |  |  | 3.89 | 5.75 |  |  |  |  |  |  | 3.89 | 5.75 | 0.11 | 0.01 |
| 8.5 |  |  | 4.06 | 7.45 |  |  |  |  |  |  | 4.06 | 7.45 | 0.11 | 0.01 |
| 9.0 |  |  | 3.07 | 6.97 |  |  |  |  |  |  | 3.07 | 6.97 | 0.09 | 0.01 |
| 9.5 |  |  | 1.64 | 4.55 |  |  |  |  |  |  | 1.64 | 4.55 | 0.05 | 0.01 |
| 10.0 |  |  | 2.35 | 7.91 |  |  |  |  |  |  | 2.35 | 7.91 | 0.07 | 0.01 |
| 105 |  |  | 1.19 | 4.79 |  |  |  |  |  |  | 1.19 | 4.79 | 0.03 | 0.01 |
| 11.0 |  |  | 1.09 | 5.36 |  |  |  |  |  |  | 1.09 | 5.36 | 0.03 | 0.01 |
| 115 |  |  | 1.31 | 7.52 |  |  |  |  |  |  | 1.31 | 7.52 | 0.04 | 0.01 |
| 12.0 |  |  | 1.08 | 7.24 | 1.35 | 9.07 |  |  |  |  | 2.43 | 1631 | 0.07 | 0.02 |
| 125 |  |  | 0.94 | 7.23 | 4.16 | 32.12 |  |  |  |  | 5.09 | 3935 | 0.14 | 0.05 |
| 13.0 |  |  | 0.66 | 5.96 | 1432 | 129.09 | 1.55 | 14.01 |  |  | 16.53 | 149.06 | 0.46 | 0.19 |
| 135 |  |  | 0.00 | 0.00 | 52.57 | 547.71 | 0.67 | 6.96 |  |  | 5324 | 554.67 | 1.50 | 0.72 |
| 14.0 |  |  | 0.00 | 0.00 | 117.50 | 140687 | 17.15 | 205.34 |  |  | 134.65 | 161222 | 3.79 | 2.08 |
| 145 |  |  | 0.00 | 0.00 | 195.36 | 266981 | 54.79 | 748.76 |  |  | 250.16 | 341858 | 7.04 | 4.42 |
| 15.0 |  |  | 0.00 | 0.00 | 284.52 | 4417.62 | 62.27 | 966.77 |  |  | 346.79 | 538439 | 9.75 | 6.96 |
| 15.5 |  |  | 0.00 | 0.00 | 225.96 | 395887 | 173.90 | 3046.67 |  |  | 399.86 | 700555 | 1125 | 9.05 |
| 16.0 |  |  | 0.00 | 0.00 | 127.08 | 2510.07 | 314.72 | 6216.44 | 6.46 | 127.57 | 448.25 | 8854.08 | 12.61 | 11.44 |
| 165 |  |  | 0.00 | 0.00 | 102.76 | 227740 | 378.36 | 8385.60 | 0.94 | 2090 | 482.061 | 068390 | 1356 | 13.81 |
| 17.0 |  |  | 0.00 | 0.00 | 5831 | 1444.71 | 353.35 | 8754.66 | 0.80 | 19.90 | 412.471 | 021927 | 11.60 | 1321 |
| 175 |  |  | 0.00 | 0.00 | 2736 | 755.79 | 324.02 | 8950.81 | 0.87 | 24.17 | 352.25 | 9730.77 | 9.91 | 1258 |
| 18.0 |  |  | 0.00 | 0.00 | 10.01 | 307.42 | 269.08 | 8259.78 | 7.07 | 217.15 | 286.17 | 878435 | 8.05 | 1135 |
| 185 |  |  | 0.00 | 0.00 | 15.63 | 531.77 | 155.58 | 529187 | 6.37 | 216.70 | 177.59 | 604033 | 5.00 | 7.81 |
| 19.0 |  |  | 0.00 | 0.00 | 0.00 | 0.00 | 79.86 | 3001.75 | 1.22 | 45.94 | 81.08 | 304770 | 2.28 | 3.94 |
| 195 |  |  | 0.00 | 0.00 | 0.00 | 0.00 | 26.65 | 1102.46 | 3.78 | 156.56 | 30.43 | 1259.02 | 0.86 | 1.63 |
| 20.0 |  |  | 0.00 | 0.00 | 0.00 | 0.00 | 8.41 | 382.18 | 0.00 | 0.00 | 8.41 | 382.18 | 0.24 | 0.49 |
| 205 |  |  | 0.00 | 0.00 | 0.00 | 0.00 | 2.00 | 99.06 | 0.08 | 3.72 | 2.07 | 102.78 | 0.06 | 0.13 |
| Sum | 42.61 | 28.89 | 25.7, | 75.61 | 1236.902 | 099833 | 2222345 | 5433.14 | 27.61 | 832.61 | 3555247 | 736858 |  |  |
| \% | 1.20 | 0.04 | 0.73 | 0.10 | 34.79 | 27.14 | 62.51 | 71.65 | 0.78 | 1.08 |  |  | 100.00 | 100.00 |

Table 9.2 Barents Sea CAPELIN. Catch statistic table. Catch 1972-2010. Thousand tonnes.


In brackets- reseach quota and catch.
*Include catch by other countries.
${ }^{* *}$ Recommended for spring season only.
*** Expert assesment catch during Russian polar cod fishery in Autumn 2009.

Table 9.3. Barents Sea CAPELIN. Stock size estimation table. Estimated stock size from the acoustic survey in August-October 2010.

| Length (cm) |  | Age/Year class, number $10{ }^{9}$ |  |  |  | $\begin{gathered} \hline \text { Sum } \\ -\left(10^{9}\right) \end{gathered}$ | Biomass$\left(10^{3} \mathrm{t}\right)$ | Mean weight (g) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1(2008) | 2(2007) | 3(2006) | 4(2005) |  |  |  |
|  | 6.0 | 0.022 |  |  |  | 0.022 | 0.0 | 1.0 |
|  | 6.5 | 0.211 |  |  |  | 0.211 | 0.2 | 1.0 |
|  | 7.0 | 0.695 |  |  |  | 0.695 | 0.8 | 1.2 |
|  | 7.5 | 1.932 |  |  |  | 1.932 | 3.3 | 1.7 |
|  | 8.0 | 9.910 | 0.068 |  |  | 9.979 | 19.8 | 2.0 |
|  | 8.5 | 19.592 |  |  |  | 19.592 | 46.9 | 2.4 |
|  | 9.0 | 22.901 | 0.808 |  |  | 23.709 | 67.1 | 2.8 |
|  | 9.5 | 24.444 | 0.360 |  |  | 24.804 | 82.3 | 3.3 |
|  | 10.0 | 25.150 | 1.673 |  |  | 26.824 | 106.1 | 4.0 |
|  | 10.5 | 11.826 | 3.286 |  |  | 15.112 | 72.4 | 4.8 |
|  | 11.0 | 4.929 | 5.089 |  |  | 10.018 | 54.7 | 5.5 |
|  | 11.5 | 1.400 | 16.871 |  |  | 18.271 | 117.7 | 6.4 |
|  | 12.0 | 0.437 | 16.257 |  |  | 16.694 | 122.5 | 7.3 |
|  | 12.5 | 0.420 | 26.085 | 0.009 |  | 26.513 | 224.1 | 8.5 |
|  | 13.0 | 0.062 | 24.569 | 0.140 |  | 24.770 | 242.4 | 9.8 |
|  | 13.5 | 0.024 | 24.103 | 0.325 |  | 24.452 | 272.3 | 11.1 |
|  | 14.0 | 0.015 | 13.755 | 1.494 |  | 15.263 | 192.9 | 12.6 |
|  | 14.5 | 0.051 | 9.958 | 1.581 |  | 11.590 | 164.5 | 14.2 |
|  | 15.0 |  | 6.002 | 3.677 |  | 9.679 | 159.1 | 16.4 |
|  | 15.5 |  | 5.313 | 6.720 |  | 12.033 | 222.8 | 18.5 |
|  | 16.0 |  | 5.065 | 8.415 |  | 13.479 | 282.5 | 21.0 |
|  | 16.5 |  | 3.255 | 12.818 | 0.029 | 16.101 | 377.3 | 23.4 |
|  | 17.0 |  | 1.151 | 9.490 | 0.061 | 10.702 | 287.2 | 26.8 |
|  | 17.5 |  | 1.996 | 7.073 | 0.161 | 9.230 | 278.6 | 30.2 |
|  | 18.0 |  | 0.650 | 6.439 |  | 7.089 | 234.5 | 33.1 |
|  | 18.5 |  | 0.068 | 1.952 |  | 2.020 | 72.2 | 35.7 |
|  | 19.0 |  |  | 1.206 |  | 1.206 | 46.9 | 38.9 |
|  | 19.5 |  |  | 0.127 |  | 0.127 | 5.5 | 43.4 |
|  | (109) | 124.021 | 166.382 | 61.465 | - 0.251 | 352.118 |  |  |
| TSB | (103 t) | 417.4 | 1821.8 | 1510.2 | -7.1 |  | 3756.5 |  |
| Mean leng | (cm) | 9.6 | 13.4 | 16.8 | 17.5 | 12.7 |  |  |
| Mean weig | (g) | 3.4 | 10.9 | 24.6 | 28.4 |  |  | 10.7 |
| SSN | (106) | 0.066 | 47.213 | 60.992 | 0.251 | 108.522 |  |  |
| SSB | $103 \mathrm{t})$ | 0.9 | 809.0 | 1505.7 | 7.2 |  | 2322.9 |  |
|  |  |  | Based on | $n$ TS value: 19 | $19.1 \log L-74$ | 4.0, correspond | ding to $\sigma=5.0$ | - $10 \cdot L 1.9$ |

Table 9.4 Barents Sea CAPELIN. Recruitment and natural mortality table. Larval abundance estimate in June, 0 -group indices and acoustic assessment in August-September, total mortality from age $1+$ to age $2+$.

| Year class | Larval 0-g roup Index ( $10^{9} \mathrm{ind}$.) Survey assesment ( $10^{9} \mathrm{ind}$.) |  |  |  |  | Z |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Abundance <br> (10 ${ }^{12}$ ) | without Keff | with Keff | $\begin{array}{r} 1+ \\ (\mathrm{Y}+1) \end{array}$ | $\begin{array}{r} 2+ \\ (\mathrm{Y}+2) \end{array}$ | $1-2 \%$ <br> (by survey) |
| 1980 | - | 197.3 | 740.3 |  |  |  |
| 1981 | 9.7 | 123.9 | 477.3 |  |  |  |
| 1982 | 9.9 | 168.1 | 599.6 | 514.9 | 186.5 | 64 |
| 1983 | 9.9 | 100.0 | 340.2 | 154.8 | 48.3 | 69 |
| 1984 | 8.2 | 68.1 | 275.2 | 38.7 | 4.7 | 88 |
| 1985 | 8.6 | 21.3 | 63.8 | 6.0 | 1.7 | 72 |
| 1986 | 0.0 | 11.4 | 41.8 | 37.6 | 28.7 | 24 |
| 1987 | 0.3 | 1.2 | 4.0 | 21.0 | 17.7 | 16 |
| 1988 | 0.3 | 19.6 | 65.1 | 189.2 | 177.6 | 6 |
| 1989 | 7.3 | 251.5 | 862.4 | 700.4 | 580.2 | 17 |
| 1990 | 13.0 | 36.5 | 115.6 | 402.1 | 196.3 | 51 |
| 1991 | 3.0 | 57.4 | 169.5 | 351.3 | 53.4 | 85 |
| 1992 | 7.3 | 1.0 | 2.3 | 2.2 | 3.4 | - |
| 1993 | 3.3 | 0.3 | 1.0 | 19.8 | 8.1 | 59 |
| 1994 | 0.1 | 5.4 | 13.9 | 7.1 | 11.5 | - |
| 1995 | 0.0 | 0.9 | 2.9 | 81.9 | 39.1 | 52 |
| 1996 | 2.4 | 44.3 | 136.7 | 98.9 | 72.6 | 27 |
| 1997 | 6.9 | 54.8 | 189.4 | 179.0 | 101.5 | 43 |
| 1998 | 14.1 | 33.8 | 113.4 | 156.0 | 110.6 | 29 |
| 1999 | 36.5 | 85.3 | 287.8 | 449.2 | 218.7 | 51 |
| 2000 | 19.1 | 39.8 | 140.8 | 113.6 | 90.8 | 20 |
| 2001 | 10.7 | 33.6 | 90.2 | 59.7 | 9.6 | 84 |
| 2002 | 22.4 | 19.4 | 67.1 | 82.4 | 24.8 | 70 |
| 2003 | 11.9 | 94.9 | 340.9 | 51.2 | 13.03 | 75 |
| 2004 | 2.5 | 16.7 | 53.9 | 26.94 | 21.7 | 19.3 |
| 2005 | 8.8 | 41.8 | 148.5 | 60.1 | 54.7 | 9.0 |
| 2006 | 17.1 | 166.4 | 515.8 | 277.2 | 231.4 | 17 |
| 2007 | - | 157.9 | 480.1 | 313.0 | 166.4 | 46 |
| 2008 | - | 288.8 | 995.1 | 124.0 |  |  |
| 2009 |  | 189.8 | 673.0 |  |  |  |
| Average | 9.0 | 77.71 | 266.9 | 167.3 | 95.1 | 45.6 |

Table 9.5 Barents Sea CAPELIN. Stock size in numbers by age, total stock biomass, biomass of the maturing component at 1 . October.

| Year | Stock in numbers (109) |  |  |  |  |  |  | Stock in weight |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 1 | Age 2 | Age 3 | Age 4 | Age 5 |  | Total | Total |  | Maturing |
|  | 1973 | 528 | 375 | 40 | 17 | 0 | 961 |  | 5144 | 1350 |
|  | 1974 | 305 | 547 | 173 | 3 | 0 | 1029 |  | 5733 | 907 |
|  | 1975 | 190 | 348 | 296 | 86 | 0 | 921 |  | 7806 | 2916 |
|  | 1976 | 211 | 233 | 163 | 77 | 12 | 696 |  | 6417 | 3200 |
|  | 1977 | 360 | 175 | 99 | 40 | 7 | 681 |  | 4796 | 2676 |
|  | 1978 | 84 | 392 | 76 | 9 | 1 | 561 |  | 4247 | 1402 |
|  | 1979 | 12 | 333 | 114 | 5 | 0 | 464 |  | 4162 | 1227 |
|  | 1980 | 270 | 196 | 155 | 33 | 0 | 654 |  | 6715 | 3913 |
|  | 1981 | 403 | 195 | 48 | 14 | 0 | 660 |  | 3895 | 1551 |
|  | 1982 | 528 | 148 | 57 | 2 | 0 | 735 |  | 3779 | 1591 |
|  | 1983 | 515 | 200 | 38 | 0 | 0 | 754 |  | 4230 | 1329 |
|  | 1984 | 155 | 187 | 48 | 3 | 0 | 393 |  | 2964 | 1208 |
|  | 1985 | 39 | 48 | 21 | 1 | 0 | 109 |  | 860 | 285 |
|  | 1986 | 6 | 5 | 3 | 0 | 0 | 14 |  | 120 | 65 |
|  | 1987 | 38 | 2 | 0 | 0 | 0 | 39 |  | 101 | 17 |
|  | 1988 | 21 | 29 | 0 | 0 | 0 | 50 |  | 428 | 200 |
|  | 1989 | 189 | 18 | 3 | 0 | 0 | 209 |  | 864 | 175 |
|  | 1990 | 700 | 178 | 16 | 0 | 0 | 894 |  | 5831 | 2617 |
|  | 1991 | 402 | 580 | 33 | 1 | 0 | 1016 |  | 7287 | 2248 |
|  | 1992 | 351 | 196 | 129 | 1 | 0 | 678 |  | 5150 | 2228 |
|  | 1993 | 2 | 53 | 17 | 2 | 2 | 75 |  | 796 | 330 |
|  | 1994 | 20 | 3 | 4 | 0 | 0 | 28 |  | 200 | 94 |
|  | 1995 | 7 | 8 | 2 | 0 | 0 | 17 |  | 193 | 118 |
|  | 1996 | 82 | 12 | 2 | 0 | 0 | 96 |  | 503 | 248 |
|  | 1997 | 99 | 39 | 2 | 0 | 0 | 140 |  | 911 | 312 |
|  | 1998 | 179 | 73 | 11 | 1 | 0 | 263 |  | 2056 | 931 |
|  | 1999 | 156 | 101 | 27 | 1 | 0 | 285 |  | 2776 | 1718 |
|  | 2000 | 449 | 111 | 34 | 1 | 0 | 595 |  | 4273 | 2099 |
|  | 2001 | 114 | 219 | 31 | 1 | 0 | 364 |  | 3630 | 2019 |
|  | 2002 | 60 | 91 | 50 | 1 | 0 | 201 |  | 2210 | 1290 |
|  | 2003 | 82 | 10 | 11 | 1 | 0 | 104 |  | 533 | 280 |
|  | 2004 | 51 | 25 | 6 | 1 | 0 | 82 |  | 628 | 294 |
|  | 2005 | 27 | 13 | 2 | 0 | 0 | 42 |  | 324 | 174 |
|  | 2006 | 60 | 22 | 6 | 0 | 0 | 88 |  | 787 | 437 |
|  | 2007 | 222 | 55 | 4 | 0 | 0 | 280 |  | 1882 | 844 |
|  | 2008 | 313 | 231 | 25 | 2 | 0 | 571 |  | 4427 | 2468 |
|  | 2009 | 124 | 166 | 61 | 0 | 0 | 352 |  | 3756 | 2323 |

Table 9.6 Barents Sea CAPELIN. Summary stock and data for prognoses table.

| Year | Estimated stock by autumn acoustic survey ( $10^{3} \mathrm{t}$ ) 1 Octo ber |  | Spawning stock biomass, assessment model, April 1 ( $10^{3} \mathrm{t}$ ) | Spawning stock biomass, by winter acoustic survey ( $10^{3} \mathrm{t}$ ) | Recruitment <br> Age 1+, survey assessment 1 Oc tober $10^{9} \mathrm{sp}$. | Young herring biomass age 1 and 2 in the Barents sea.$\left(10^{3} t\right)$ | Landing$\left(10^{3} \mathrm{t}\right)$ | Rate of the TSB increase |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | TSB | SSB |  |  |  |  |  |  |
| 1972 | 6600 | 2727 |  |  |  |  | 1592 |  |
| 1973 | 5144 | 1350 | 33 |  | 528 | 2 | 1336 | 0.8 |
| 1974 | 5733 | 907 | * |  | 305 | 48 | 1149 | 1.1 |
| 1975 | 7806 | 2916 | * |  | 190 | 74 | 1440 | 1.4 |
| 1976 | 6417 | 3200 | 253 |  | 211 | 39 | 2587 | 0.8 |
| 1977 | 4796 | 2676 | 22 |  | 360 | 46 | 2987 | 0.7 |
| 1978 | 4247 | 1402 | * |  | 84 | 52 | 1915 | 0.9 |
| 1979 | 4162 | 1227 | * |  | 12 | 39 | 1783 | 1.0 |
| 1980 | 6715 | 3913 | * |  | 270 | 66 | 1648 | 1.6 |
| 1981 | 3895 | 1551 | 316 |  | 403 | 47 | 1986 | 0.6 |
| 1982 | 3779 | 1591 | 106 |  | 528 | 9 | 1760 | 1.0 |
| 1983 | 4230 | 1329 | 100 |  | 514,9 | 12 | 2358 | 1.1 |
| 1984 | 2964 | 1208 | 109 |  | 154,8 | 1313 | 1478 | 0.7 |
| 1985 | 860 | 285 | * |  | 38,7 | 1220 | 868 | 0.3 |
| 1986 | 120 | 65 | * |  | 6,0 | 155 | 123 | 0.1 |
| 1987 | 101 | 17 | 34 | 3,63 | 37,6 | 145 | 0 | 0.8 |
| 1988 | 428 | 200 | * | 10,3 | 21,0 | 70 | 0 | 4.2 |
| 1989 | 864 | 175 | 84 | 378,09 | 189,2 | 126 | 0 | 2.0 |
| 1990 | 5831 | 2617 | 92 | 94,2 | 700,4 | 356 | 0 | 6.7 |
| 1991 | 7287 | 2248 | 643 | 1769,7 | 402,1 | 646 | 933 | 1.2 |
| 1992 | 5150 | 2228 | 302 | 1734,8 | 351,3 | 1537 | 1123 | 0.7 |
| 1993 | 796 | 330 | 293 | 1498,39 | 2,2 | 2466 | 586 | 0.2 |
| 1994 | 200 | 94 | 139 | 187,4 | 19,8 | 1715 | 0 | 0.3 |
| 1995 | 193 | 118 | 60 | 29,83 | 7,1 | 558 | 0 | 1.0 |
| 1996 | 503 | 248 | 60 |  | 81,9 | 208 | 0 | 2.6 |
| 1997 | 909 | 312 | 85 |  | 98,9 | 273 | 0,5 | 1.8 |
| 1998 | 2056 | 932 | 94 | 413,59 | 179,0 | 376 | 3,02 | 2.3 |
| 1999 | 2775 | 1718 | 382 |  | 156,0 | 1201 | 104 | 1.3 |
| 2000 | 4273 | 2098 | 599 | 699,9 | 449,2 | 1766 | 410 | 1.5 |
| 2001 | 3630 | 2019 | 626 |  | 113,6 | 949 | 577,6 | 0.8 |
| 2002 | 2210 | 1291 | 496 | 1416,88 | 59,7 | 309 | 660,3 | 0.6 |
| 2003 | 533 | 280 | 427 |  | 82,4 | 2197 | 281,54 | 0.2 |
| 2004 | 628 | 294 | 94 | 104,94 | 51,2 | 2556 | 0 | 1.2 |
| 2005 | 324 | 174 | 122 |  | 26,94 | 1878 | 1,21 | 0.5 |
| 2006 | 787 | 437 | 72 |  | 60,1 | 1335 | 0 | 2.4 |
| 2007 | 2119 | 844 | 189 |  | 277,2 | 408 | 4,0 | 2.7 |
| 2008 | 4428 | 2468 | - | 468,9 | 313,0 | 232 | 12,0 | 2.1 |
| 2009 | 3765 | 2323 | 517 | 180,03 | 124.02 | 60 | 306,14 | 0.9 |
| 2010 |  |  | 504 | 451.9 |  |  |  |  |



Figure 9.1. Geographical distribution of capelin during the acoustic survey in autumn 2009 (1:0-25, 2: 25-50, 3: 50-75, 4: > 75 t/nm $\mathbf{n m}^{2}$ )


Figure 9.2. Probabilistic prognos is 1 October 2009-1 April 2010 for Barents Sea capelin (maturing stock, catch of 360000 tonnes).


Figure 9.3. Probability of spawning biomass of capelin (1 April 2010) being below $B_{\text {lim }}$ (200 000 tonnes), as a function of catch.


Figure 9.4. Regression of abundance of capelin at age 0 ( 0 -group index without $K_{\text {eff }}$ ) and age 1 (acoustic estimate) of year classes 1981-2008. The regression line is forced through the origin, to avoid systematic overestimation of weak year classes.


Figure 9.5. Spawning stock-recruitment plot for capelin, with colours of points indicating different levels of young herring abundance.

## 10 Working documents

## WD\# Title

1 Protocol of an international worksho p on c apelin o tolith reading
2 Capelin spawning stock, prespawning migrations and Russian fishery in winter-spring 2010
3 The Spanish NE Arctic Cod Fishery in 2009
4 Consumption of various prey species by cod in 1984-2009
5 Retrieving the times series of input data for assessment of NEA saithe
6 Data availability and critical gaps in knowle dge in estimation of Catch at age for 3 stocks in the Norwegian Northeast Arctic fishery.

7 Reporton the Portuguese fishery 2009:ICES Div. I, IIa, Ilb

8 Acoustic abundance of saithe, coastal cod and haddock Finnmark - Møre, autumn 2009
9 Results of the Russian survey of Greenland halibut in the Barents Sea and adjacent waters in 2009
10 Methods for estimating F for coastal cod
11 Spatial resolution of German CPUE for Northeast Arctic saithe (Pollachius Bernreuther, Fock and virens L.) in ICES Div ision IIa from 1995 to 2009

12 Evaluation of haddock predation on the Barents Sea capelin stock
13 Estimation of to tal catches of cod and haddock in the Barents Sea in 2009 incl. Comparison of vessel information

14 Report of the 2009 meeting between the Norwegian and Russian age reading specialists

15 Results from the Joint IMR-PINRO Barents Sea demersal fish survey 1 February - 17 March 2010
16 Spanish bottom trawl May survey fletan artico 2009 in sthe slope of Svalbard area, ICES division IIb.
Data series on recreational and tourist fisheries for Norwegian Coastal Cod Sunnanå
17 Data series on recreational and tourist fisheries for Norwegian Coastal Cod Sunnanå
18 Current XS A analy sis on Greenland halibut is sensitive to changes in some Hallfredsson model assumptions
19 Howcan we assess recruitment models for (age-3) NEA cod? Dingsør et al.
20 Indices of abundance from the Joint Norwegian-Russian EcosystemSurvey Prozorkevitch in autumn

21 Stochastic Approach for Evaluating of Capelin Impacton NEA CodS tock Filin Dynamics
22 Assessment of population recruitment abundance of Northeast Arctic cod Titov considering the environment data
23 The current situation of climate, phy to plankton, zoo plankton, shrimp, harp seal and fish in the Barents Sea 2009 and beginning of 2010
24 An assessment of the future assessment site Stiansenet al.
25 To problem on methodology of Greenland halibut age reading by different Kuznetsova registering structures

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## Standard Procedure for Assessment

## XSA/ICA Type

Stock specific documentation of standard assessment procedures used by ICES.

Stock: Norwegian Coastal cod<br>Working Group: Arctic Fisheries Working Group<br>Date:<br>11-05-2010

## Approach used by the 2010 WG

For several years the xsa-analyses based on this stock annex have shown a retrospective bias. At the same time the trends seen in the survey and the catches have been considered to be a sufficient basis for the advice. The 2010 wg was asked to evaluate a rebuilding plan for coastal cod. It was then a need for a more robust analytical assessment. In addition, a new time series on catch at age in the recreational fishery was presented and added to the canum for commercial catches.

An estimate for F 2009 was obtained from surveys and an estimate for F2008 were obtained directly from catches (details in Annex 10). These estimates were used for deciding on a best estimate of F2009 that were used as terminal F in a traditional vpa. Selection at age in 2009 and Fold for earlier years were taken from a trial xsa. In addition to this, the annual values for maturity were replaced by the average observed over the survey series (1995-2009).

The traditional vpa were then taken as the final assessment.
With the new catch data the xsa showed improved diagnostics, particularly for the younger ages, when assuming catchability dependent on stock numbers for ages 2 and 3.

Some of these changes were rather ad hoc. Some intercessional further work should examine this further, and a benchmark would be relevant in near future.

Chapters A-I is the stock Annex dated 24. April 2009.

## A General

## A.1. Stock definition

Cod in the Barents Sea, the Norwegian Sea and in the coastal areas living under variable environmental conditions form groups with some peculiarities in geographical distribution, migration pattern, growth, maturation rates, genetics features, etc. The degree of intermingle of different groups is uncertain (Borisov, Ponomarenko and Yaragina, 1999). However, taking into account some biological characteristics of cod in the coastal zone and the specifics of the coastal fishery, the Working Group consi-
dered it acceptable to assess the Norwegian coastal cod stock (in the frame of ICES) separately from North-East Arctic cod.

Both types of cod (the Norw egian Coastal cod and the North-East Arctic cod) can be met together on spawning grounds during spawning period as well as in catches all the year round both inshore and offshore in variable proportions.

The Norwegian Coastal cod (NCC) is distributed in the fjords and along the coast of Norway from the Kola peninsula in northeast and south to Møre at $62^{\circ}$ N. Spawning areas are located in fjords as well as offshore along the coast. Spawning season extents from March to late June. The 0 and 1-group of NCC inhabit shallow water both in fjords and in coastal areas and are hardly found in deeper trawling areas until reaching about 25 cm . Afterwards they gradually move towards deeper water. NCC starts on average to mature at age 4-6 and migrates towards spawning grounds in early winter. The majority of the biomass (about $75 \%$ ) is located in the northern part of the area (North of $67^{\circ} \mathrm{N}$ ).

Tagging experiments of cod inhabiting fjords indicate only short migrations (Jakobsen 1987, Nøstvik and Pedersen 1999, Skreslet, et al. 1999). From these experiments very few tagged cod migrated into the Barents Sea ( $<1 \%$ ). Investigations based on genetics find large difference between NCC and North-East Arctic cod (NEAC) (Fevolden and Pogson 1995, Fevolden and Pogson 1997, Jørstad and Nævdal 1989, Møller 1969), while others do not find clear differences (Árnason and Pálsson 1996, Mork, et al. 1984, Artemjeva and Novikov, 1990). Investigations also indicate that NCC probably consists of several separate populations.

Ongoing microsatellite studies on the genetic structure of cod along the entire Norwegian coast have revealed considerable genetic differences. Two main clusters were indicated: one north of 64 deg north (Trondheimsfjord) and one to the south of this. Differences were also observed between regions within these clusters. The conclusion is that NCC is not a single stock.

## A.2. Fishery

Coastal cod is mainly fished by small coastal vessels using traditional fishing gears like gillnet, longline, hand line and danish seine, but some is also fished by trawlers and larger longliners fishing at the coastal banks. The fishery is dominated by gillnet (50\%), while longline/hand line account for about $20 \%$, Danish seine $20 \%$ and Trawl $10 \%$ of the total catch. There was a shift around 1995 in the portion caught by the different gears. Before 1995 the portion taken by longline and hand line was higher, while the portion taken by danish seine was lower. Norwegian vessels take all the reported catch. How ever, trawlers from other countries probably take a small amount of NCC when fishing near the Norwegian coast fishing for North-East Arctic cod and North-East Arctichaddock.

The TAC set for coastal cod is added to the Norwegian TAC for North-east Arctic cod, giving a total, combined TAC to distribute on fishing vesslels. Cod catches are not identified to stock at landing, and therefore no landings are counted against a separate coastal cod quota. When the fishing year is finished the catches of coastal cod are estimated from otholit sampling. All regulations for North-east Arctic cod also applies to coastal cod. This includes minimum catch size, minimum mesh size , maximum by-catch of undersized fish, and closure of areas having high densities of juveniles. In addition, trawl fishing for cod is not allowed inside the 6-n.mile, and since the mid 90-ies the fords in Finnmark and northern Troms (areas 03 and 04) has been closed for fishing with Danish seine, and since 2000 the large longliners have
been given restrictions, now only allowed to fish outside the 4 n.mile. Since 2004 additional restrictions on coastal fisheries have been introduced to reduce catches of coastal cod. In these new regulations "fjord-lines" are drawn along the coast to close the fjords for direct cod fishing with vessels larger than 15 meter. Abox closed for all fishing gears except hand-line and fishing rod is defined in the Henningsvær-Svolvær area. This is an area where spawning concentrations of coastal cod is usually observed and where the catches of coastal cod has been high. Since the coastal cod is fished under a combined coastal cod/north-east arctic cod quota, these regulations are supposed to turn parts of the traditional coastal fishery over from catching coastal cod in the fjords to catch more cod outside the fjords where the proportion of Northeast Arctic cod is higher. Further restrictions were introduced in 2007 by not allowing pelagic gill net fishing for cod and by reducing the allowed by-catch of cod when fishing for other species inside fjord lines from $25 \%$ to $5 \%$, and outside fjord-lines from $25 \%$ to $20 \%$. In 2009 a fjord area off Ålesund was closed in the spawning season for fishing with all gears except handline and fishing rod.

## A.3. Ecosystem aspects

Not investigated

## B. Data

## B. 1 Commercial catch

From 1996, cod caught inside the 12 n.mile zone have been separated into Norwegian coastal cod and North-east Arctic cod based on biological sampling (Berg, et al. 1998) The method is based on otolith-typing. This is the same method as is used in separating the two stocks in the surveys targeting NEAC. The catches of Norwegian coastal cod (NCC) have been calculated back to 1984using available data on otolith typing. During this period the catches have been between 22,000 and $75,000 \mathrm{t}$.

The separation of the Norwegian catches into NEAC and NCC is based on:

- No catches outside the 12 n.mile zone have been allocated to the NCC catches.
- The catches inside 12 n.mile zone are separated into quarter, fishing gear and Norw egian statistical areas.
- From the otolith structure, catches inside the 12 n.mile zone have been allocated to NCC and NEAC. The Institute of Marine Research in Bergen has been taking samples of commercial catches along the coast for a long period.

Norwegian commercial catch in tonnes by quarter, area and gear are derived from the sales notes statistics of The Directorate of Fisheries. Data from 8 sub areas are aggregated on 6 main areas for the gears gillnet, long line, hand line, Danish seine and trawl. No discards are reported or accounted for, but there are reports of discards and incorrect landings with respect to fish species and amount of catch. The scientific sampling strategy from the commercial fishing is to have age-length samples from all major gears in each area and quarter. The sampling intensity is determined by knowledge on the distribution of the combined cod catches.
There are at present no defined criteria on how to allocate samples of catch numbers, mean length and mean weight at age to unsampled catches. The following general process has been applied: First look for samples from a neighbouring area if the fishery extends to this area in the same quarter. If there are no samples available in
neighbouring areas, search for samples from other gears with the most similar selectivity in the same area or in neighbouring areas. The last option is to search in neighbouring quarters, first from the same gear in the same area, and than from neighbouring areas and similar gears. Age-length keys from research surveys with shrimp trawl (Norwegian coastal survey) are also used to fill holes.

Weight at age is calculated from the commercial catch back to 1984. The mean values are weighted by catches in the respective areas.

Proportions mature at age from 1984 to 1994 are obtained from the commercial catch data. From 1995 onwards the proportions mature at age are obtained from the Norwegian coastal survey.
Norway is assumed to account for all NCC landings. The text table below shows which kind of data are collected:

|  | KIND OF DATA |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Country | Caton (catch <br> in weight) | Canum (catch <br> at age in <br> numbers) | Weca(weight <br> atage in the <br> catch) | Matprop <br> (proportion <br> mature by <br> age) | Length <br> composition <br> in catch |
| Norway | X | X | X | X | X |

## B.2. Biological

Weight at age in the stock is obtained from the Norwegian coastal survey in from 1995 onwards. From 1984 to 1994 w eight at age in stock is taken from weight at age in the catch because no survey data from this period are available. The mean values are weighted by biomass in the respective areas. In 2007 a weight at age series of unweighted mean values from the survey was calculated and used in the SURBA analysis.

A fixed natural mortality of 0.2 is used both in the assessment and the forecast. Some fjord studies (Pedersen and Pope, 2003a and b, Mortensen 2007, Pedersen et al., 2007). indicate that the main predators on young cod is larger cod, cormorants and saithe. There are no estimates of annual predation mortality for the stock complex.
Both the proportion of natural mortality before spawning (Mprop) and the proportion of fishing moratlity before spawning (Fprop) are to 0 .

## B.3. Survey

Since 1995 a Norwegian trawl-acoustic survey (Norwegian coastal survey) specially designed for coastal cod has been conducted annually in September (prior to 2003) and in October-November ( 28 days). The survey covers the fjords and coastal areas from the Varangerfjord close to the Russian border and southwards to $62^{\circ} \mathrm{N}$. The aim of conducting a acoustic survey targeting Norwegian coastal cod has been to support the stock assessment with fishery-independent data of the abundance of both the commercial size cod as well as the youngest pre-recruit coastal cod. The survey therefore covers the main areas where the commercial fishery takes place, normally dominated by 4-7 year old fish.

The 0 - and 1 year-old coastal cod, mainly inhabiting shallow water ( $0-50$ meter) near the coast and in the fords, are also represented in the survey, although highly variable from year to year. However, the 0-group cod caught in the survey is impossible to classify to NCC or NEAC by the otoliths since the first winter zone is used in this
separation. A total number of more than 200 trawl hauls are conducted during the survey ( 100 bottom trawl, 100 pelagic trawl).

The survey abundance indexes at age are total numbers (in thousands) computed from the acoustics.

Ages 2-8 are used in the XSA-tuning. Ages $2-9$ are used in a SURBA analysis.

## B.4. Commercial CPUE

No commercial CPUE are available for this stock.

## B.5. Other relevant data

A number of bottom trawl tows are made during the coastal survey, and since 2003 the survey has aimed for towing at the same fixed positions each year. This might be used to calculate a bottom trawl index.

## C. Historical stock development

## Acoustic survey

The total acoustic biomass varies between $144,000 \mathrm{t}$ (1995) and 30,300t (2005), showing a decline from 1995 until 2003, and flat level since 2003. The indices show considerable year to year variations. The acoustic spawning biomass vary between $75,000 t$ (1995) and 12,700t (2005), showing the same type of trend as the total biomass. The recruitment of 2 year old fish vary from 20 million individuals in 1995 to 2 million in 2005, also showing the same, but stronger trend as the total stock.

## SURBA analysis

The SURBA analysis (SURBA 2.10) is run with the same data as input to the XSA (se below). However, the age span is $2-9$ year in the SURBA analysis. The settings are set similar to the XSA settings. The weight at age for the stock is calculated as unweighted mean values to avoid some of the large fluctuations in the weight at age from the survey calculations.

The history of the stock is reflected in the same way in this analysis as in the survey, showing a drop to a level in the later years about $25 \%$ of the level in 1995. The recruitment is down to a $10 \%$ level.

## VPA analysis

Model used: XSA
Software used: IFAP / Low estoft VPA suite
Model Options chosen:
Tapered time weighting applied, power $=3$ over 20 years
Catchability independent of stock size for all ages
Catchability independent of age for ages $>=8$
Survivor estimates shrunk towards the mean F of the final 2 years or the 4 oldest ages
S.E. of the mean to which the estimate are shrunk $=1.0$

Minimum standard error for population estimates derived from each fleet $=0.300$
Prior weighting not applied

Input data types and characteristics:

| Type | Name | Year range | Age range | Variable from year to year Yes/No |
| :---: | :---: | :---: | :---: | :---: |
| Caton | Catch in tonnes | 1984 - last data year | 2-10+ | Yes |
| Canum | Catchatage in numbers | 1984 - last data year | 2-10+ | Yes |
| Weca | Weight at age in the commercial catch | 1984 - last data year | 2-10+ | Yes |
| West | Weight at age of the spawning stock at spawning time. | 1984 - last data year | 2-10+ | Yes/No - assumed to be the same as weightatage in the catch from 1984-1994 |
| Mprop | Proportion of natural mortality before spawning | 1984 - last data year | 2-10+ | No - set to 0 for all ages in all years |
| Fprop | Proportion of fishing mortality before spawning | 1984 - last data year | 2-10+ | No - set to 0 for all ages in all years |
| Matprop | Proportion mature at age | $1984 \text { - last data }$ <br> year | 2-10+ | Yes |
| Natmor | Natural mortality | 1984 - last data year | 2-10+ | No - set to 0.2 for allages in all years |
| Tuning fleet | Norwe gian coastal survey | $1995 \text { - last data }$ <br> year | 2-8 |  |

The results show a variation of the total biomass between $310,000 \mathrm{t}$ (1984) and $87,000 \mathrm{t}$ (2008) with the value in 1995 being 260,000t. The spawning stock is estimated to 170,000 t in 1995 , falling to $50,000 \mathrm{t}$ in 2008 . The fishing mortality is estimated to 0.38 on average. The pattern of stock decline is fairly similar to that of the survey.

## D. Short-term projection

No quantative projection but trends in stock biomass, mortality and recruitment obtained from surba (and xsa) are used to indicate stock development.t

## E. Medium-term projections

Not done.
F. Long-term projections

Not done.

## G. Biological reference points

Not available.

## H. Other issues

## I. References

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Stock specific documentation of standard assessment procedures used by ICES.

Stock:<br>North-East Arctic Cod<br>Working Group: Arctic Fisheries Working Group (AFWG)<br>Date:<br>27. April 2009.

## A. General

## A. 1 Stock definition

The North-East Arctic cod (Gadus morhua) is distributed in the Barents Sea and adjacent waters, mainly in waters above $0^{\circ}$ Celsius. The main spawning areas are along the Norwegian coast between $\mathrm{N} 67^{\circ} 30^{\prime}$ and $70^{\circ}$. The 0 -group cod drifts from the spawning grounds eastwards and northwards and during the international 0 -group survey in August it is observed over wide areas in the Barents Sea.

## A. 2 Fishery

The fishery for North-east Arctic cod is conducted both by an international trawler fleet operating in offshore waters and by vessels using gillnets, longlines, handlines and Danish seine operating both offshore and in the coastal areas. $60-80 \%$ of the annual landings are from trawlers. Catch quotas were introduced in the trawl fishery in 1978 and for the fisheries with conventional gears in 1989. In addition to quotas the fisheries are regulated by mesh size limitations including sorting grids, a minimum catching size, a maximum by-catch of undersized fish, maximum by-catch of nontarget species, closure of areas with high densities of juveniles and by seasonal and area restrictions. Since January 1997 sorting grids have been mandatory for the trawl fisheries in most of the Barents Sea and Svalbard area. Discarding is prohibited. The minimum catching size of cod is 42 cm in the Russian Economic zone, 47 cm in Norwegian Economic zone; both minimum landing sizes are used by respective fleets in the Svalbard area pursuant to the Svalbard Treaty 1920. The fisheries are controlled by inspections at sea, requirement of reporting to catch control points when entering and leaving the EEZs and by inspections when landing the fish for all fishing vessels. Keeping a detailed fishing log-book on board is mandatory for most vessels, and large parts of the fleet report to the authorities on a daily basis. There is some evidence that the present catch control and reporting systems are not sufficient to prevent discarding and under-reporting of catches, but it has considerably improved in comparison with historical period.

## A. 3 Ecosystem aspects

Considerable effort has been devoted to investigate multispecies interactions in the Northeast Arctic. Some of these investigations have reached the stage where quantitative results are available for use in assessments. Growth of cod depends on availability of prey such as capelin (Mallotus villosus), and variability in cod growth has had
major impacts on the cod fishery. Cod are able to compensate only partially for low capelin abundance, by switching to other prey species. This may lead to periods of high cannibalism on young cod, and may result in impacts on other prey species which are greater than those estimated for periods when capelin is abundant. In a situation with low capelin abundance, juvenile herring (Clupea harengus) experience increased predation mortality by cod. The timing of cod spawning migrations is influenced by the presence of spawning herring in the relevant area. The interaction between capelin and herring is illustrated by the recruitment failure of capelin coinciding with years of high abundance of young herring in the Barents Sea. Herring predation on capelin larvae is believed to be partially responsible for the recruitment failure of capelin when young herring are abundant in the Barents Sea.

The composition and distribution of species in the Barents Sea depend considerably on the position of the polar front which separates warm and salty Atlantic waters from colder and fresher waters of arctic origin. Variation in the recruitment of some species including cod and capelin has been associated with the changes in the influx of Atlantic waters to the large areas of the Barents Sea shelf.

The annual consumption of herring, capelin and cod by marine mammals (mainly harp seals and minke whales) has been estimated to be in the order of 1.5-2.0 million $t$ (Bogstad, Haug and Mehl, 2000; See also Section 1.3.4 AFWG Report 2003).

However, estimates of total annual food consumption of Barents Sea harp seals are in the range of about 3.3-5 million tons (depending on choice of input parameters, ICES 2000d). The applied model used different values for the field metabolic rate of the seals (corresponding to two or three times their predicted basal metabolic rate) and under two scenarios: with an abundant capelin stock and with a very low capelin stock.

1 ) If capelin was abundant the total harp seal consumption was estimated to be about 3.3 million tons (using lowest field metabolic rate). The estimated consumption of various commercially important species was as follows (in tons): capelin approximately 800,000 , polar cod (Boreogadus saida) 600,000, herring 200,000 and Atlantic cod 100,000.
2 ) A low capelin stock in the Barents Sea (as it was in 1993-1996) led to switches in seal diet composition, with estimated increased consumption of polar cod ( 870,000 tons), other codfishes (mainly Atlantic cod; 360,000 tons), and herring (390,000 tons).

## B. Data

## B. 1 Commercial catch

## Norway

Norwegian commercial catch in tonnes by quarter, area and gear are derived from the sales notes statistics of The Directorate of Fisheries. Data from about 20 sub areas are aggregated on 6 main areas for the gears gill net, long line, hand line, purse seine, Danish seine, bottom trawl, shrimp trawl and trap. For bottom trawl the quarterly area distribution of the catches is adjusted by logbook data from The Directorate of Fisheries and the total bottom trawl catch by quarter and area is adjusted so that the total annual catch for all gears is the same as the official total catch reported to ICES.
No discards are reported or accounted for, but there are several reports of discards.

The sampling strategy is to have age and length samples from all major gears in each main area and quarter. The main sampling program is sampling the landings. Additional samples from catches are obtained from the IMR reference fleet (fishing vessels contracted for sampling), and the coast guard.
A software ("ECA", Hirst et al. 2005) has been developed to utilize all sampling information to estimate catch at age for areas ( $\mathrm{I}, \mathrm{IIa}$ and IIb ), quarters and gears (bottom trawl, gill net, Danish seine and longline/handline).

## Russia

Russian commercial catch in tonnes by quarter and area are derived from the AllRussian Institute of fishery and oceanography (Moscow) statistics department. Data from each fishing vessel are aggregated on three ICES sub-Division (I, IIa and IIb).Russian fishery by passive gears was almost stopped by the end of the 1940s. At present bottom trawl fishery constitutes more than $95 \%$ cod catch.
The sampling strategy was to conduct mass measurements and collect age samples directly at sea, onboard of both research and commercial vessels to have age and length distributions from each area and quarter. Data on length distribution of cod in catches were collected in areas of cod fishery all the year round by a "standard" fishery trawl (mesh size is 125 mm in the Russian Economic zone and Svalbard area and 135 mm in the Norwegian Economic zone) and summarized by three ICES sub-areas (1, IIa and IIb). Previously the PINRO area divisions were used, differed from the ICES sub-Divisions.

Age sampling was carried out by two ways: without any selection (otoliths were taken from any fish caught in one trawl, usually from 100-300 sp.) or using a stratified by length sampling method (i.e. approximately $10-15 \mathrm{sp}$. per each $10-\mathrm{cm}$ length group). The last method has been used since 1988.

All fish taken for agereading were measured and weighted individually.
Catch at age are reported to ICES AFWG by sub-Division (1, IIa and IIb) and quarter (before 1984 - by sub-Division and year). Data on length distribution of cod in catches, as well as age-length keys, are formed for each quarter and area. In the case when a catch is present in the area/quarter but a length frequency is absent, a length frequency for the corresponding quarter, summarised for the whole sea is used. If there is no data on length composition of cod in catches per a quarter within the whole sea, a frequency summarised for the whole year and whole sea is used. Gaps in age-length distributions in sub-Divisions are filled in with data from the corresponding quarter, summarised for the whole sea. Rest gaps are filled in with information from the agelength key formed for the long-term period (1984-1997) for each quarter and for the whole sea. (Kovalev and Yaragina, 1999). Before 1984 calculation of annually catch cod numbers in sub-Divisions was derived from summarized for both the whole year age-length keys and length distribution in catches.

## Germany, Poland and Spain

Catch at age reported to the WG by ICES sub-Division (I, IIa and IIb) and quarter, according to national sampling. Missing quarters/sub-Divisions filled in by use of Russian or Norwegian sampling data.

## Other nations

Total annual catch in tonnes is reported by ICES sub-Divisions. All caches by other nations are taken by trawl. The age composition from the sampled trawl fleets is therefore applied to the catches by other nations.

The text table below shows which country supplied which kind of data for 2008:

|  | Kind of data |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Country | Caton (catch in weight) | Canum (catch at age in numbers) | Weca (weight at age in the catch) | Matprop (proportion mature by age) | Length composition in catch |
| Norway | x | x | x | x | x |
| Russia | x | x | x | x | x |
| Germany | x | x | x |  | x |
| United | x |  |  |  |  |
| King dom | x |  |  |  |  |
| France ${ }^{1}$ | x |  |  |  |  |
| Spain | x |  |  |  |  |
| Portugal | x |  |  |  | x |
| Poland | x | x | x |  |  |
| Ireland ${ }^{1}$ | x |  |  |  |  |
| Greenland ${ }^{1}$ | x |  |  |  |  |
| Faroe Islands ${ }^{1}$ | x |  |  |  |  |
| Iceland ${ }^{1}$ | x |  |  |  |  |

${ }^{1}$ As reported to Norwegian and Russian authorities
Since 2008 the catch data has been handled by Intercatch. Earlier the nations that sample the catches, provided the catch at age data and mean weights at age on Excel spreadsheet files, and the national catches were combined in Excel spreadsheet files. Historic data should be found in the national laboratories and with the stock coordinator.

For 1983 and later years mean weight at age in the catch is calculated as the weighted average for the sampled catches. For the earlier period (1946-1982) mean weight at age in catches is set equal to mean weight at age in the stock (ICES 2001).

Since 2008 the catch data has been handled by Intercatch.

## B. 2 Biological

For 1983 and later years w eight at age in the stock and maturity at age is calculated as weighted averages from Russian and Norwegian surveys during the winter season. Stock weights at age a $\left(\mathrm{W}_{\mathrm{a}}\right)$ at the start of year y are calculated as follows:

$$
W_{a}=0.5\left(W_{r u s, a-1}+\left(\frac{N_{n b a r, a} W_{\text {nbar }, a}+N_{\text {lof }, a} W_{l o f ~}, a}{}\right)\right)
$$

where
$W_{r u s a-1}$ : Weight at age a-1 in the Russian survey in year y-1
$N_{n b a r a}$ : Abundance at age a in the Norwegian Barents Sea acoustic survey in year y
$W_{\text {rbara }}$ : Weight at age a in the Norwegian Barents Sea acoustic survey in year y
$N_{l o f a}$ : Abundance at age a in the Lofoten survey in year y
$W_{b f a}$ : Weight at age a in the Lofoten survey in year y
Maturity at age is estimated from the same surveys by the same formulae, replacing weight by proportion mature.

For age groups 12 and older, the stock $w$ eights is set equal to the catch weights, since most of this fish is taken during the spawning fisheries, and in most years considerably more fish from these ages are sampled from the catches than from the surveys.

For the earlier period (1946-1982) the maturity at age and weight at age in the stock is based on Russian sampling in late autumn (both from fisheries and from surveys) and Norwegian sampling in the Lofoten spawning fishery. These data were introduced and described in the 2001 assessment report (ICES 2001).
A fixed natural mortality of 0.2 is used both in the assessment and the forecast.
Both the proportion of natural mortality before spawning (Mprop) and the proportion of fishing mortality before spawning (Fprop) are set to 0 . The peak spawning in the Lofoten area occurs most years in late March-early April.

## B. 3 Surveys

## Russia

Russian surveys of cod in the southern Barents Sea started in the late 1940s as trawl surveys of young demersal fishes. Since 1957 such surveys have been conducted over the whole feeding area including the Bear Island - Spitbergen area (Baranenkova, 1964; Trambachev, 1981), both young and adult cod have been surveyed simultaneously. In 1984, acoustic methods started to be implemented during surveys of fish stocks (Zaferman, Serebrov, 1984; Lepesevich, Shevelev, 1997; Lepesevich et al., 1999). In 1995 a new acoustic assessment method was applied for the first time, which allowed the differentiation and registration of echo intensities from fish of different length (Shevelev et al., 1998). Methods of calculations of survey indices also changed, e.g. due to the necessity to derive length-based indices for the FLEKSIBEST model (Bogstad et al.1999; Gusev, Yaragina, 2000).

Time of survey conducting has reduced from 5-6 months (September-February) in 1946-1981 to 2-2.5 months (October-December) since 1982. The aim of conducting a survey is to investigate both the commercial size cod as well as the young cod and to receive reliable data to compose annual maturity ogives. The survey covers the main areas where fries settle down as well as the commercial fishery takes place, included cod at age $0+-10+$ years. A total number of more than 400 trawl hauls are conducted during the survey (mainly bottom trawl, a few pelagic trawl).

There are two survey abundance indices at age: 1). absolute numbers (in thousands) computed from the acoustics and 2). trawl swept area indices, calculated as absolute numbers registered in survey standard area (Golovanov et al., 2006, 2007).

Ages 3-9 are used in the XSA-tuning.
Joint Russian-Norwegian winter (February) survey
The survey started in 1981 and covers the ice-free part of the Barents see. Both swept area estimates from bottom trawl and acoustic estimates are produced. The swept area estimates are used in the tuning for ages 3-8, and the acoustic estimate are added
to the Norwegian acoustic survey in Lofoten and used for tuning for ages 3-9. The survey is described in Jakobsen et al (1997) and Aglen et al. (2002).

## Norwegian Lofoten survey

Acoustic estimates from the Lofoten survey extends back to 1984. The survey is described by Korsbrekke (1997).

## B. 4 Commercial CPUE

## Russia

Two CPUE data series exist, one is historical series, based on RT vessel type (side trawler, 800-1000 HP), which stopped operating in the Barents Sea in the middle of the $1970-\mathrm{s}$, and other one is presently used, based on PST vessel type (stern trawler, 2000 HP ). Information from each fishing trawler was daily transferred to PINRO, including data on each haul (timing, location, gear and catch by species). Yearly catch $f$ cod by the PST trawlers as well as number of hour trawling were summarized and CPUE index (catch on tons per hour fishing) was calculated.
The effort (hours trawling) was scaled to the whole Russian catch. The CPUE indices are split on age groups by age data from the trawl fishery. Data on ages 9-11 are used in the XSA-tuning.

## C. Estimation of historical stock development

Model used: XSA
Software used: IFAP / Lowestoft VPA suite
Model Options chosen:
Tapered time weighting applied, power $=3$ over 10 years
Catchability independent of stock size for ages $>6$
Catchability independent of age for ages $>=10$
Survivor estimates shrunk towards the mean $F$ of the final 5 years or the 2 oldest ages
S.E. of the mean to which the estimate are shrunk $=1.000$

Minimum standard error for population estimates derived from each fleet $=0.300$
Prior weighting not applied

Input data types and characteristics:

| TYPE | Name | Year range | AGe range | VARIABLE FROM year to year Yes/No |
| :---: | :---: | :---: | :---: | :---: |
| Caton | Catch in tonnes | 1946 - last data year | 3-13+ | Yes |
| Canum | Catch at age in numbers | 1946 - last data year | 3-13+ | Yes |
| Weca | Weight at age in the commercial catch | $\begin{aligned} & 1982 \text { - last data } \\ & \text { year } \end{aligned}$ | 3-13+ | Yes, setequal to west for 19461981 |
| West | Weight at age of the spawning stock at spawning time. | 1946 - last data year | 3-13+ | Yes |
| Mprop | Proportion of natural mortality before spawning | 1946 - last data year | 3-13+ | No - set to 0 for all ages in all years |
| Fprop | Proportion of fishing mortality before spawning | $\begin{aligned} & 1960 \text { - last data } \\ & \text { year } \end{aligned}$ | 3-13+ | No - set to 0 for all ages in all years |
| Matprop | Proportion mature at age | 1960 - last data year | 3-13+ | yes |
| Natmor | Natural mortality | $\begin{aligned} & 1960 \text { - last data } \\ & \text { year } \end{aligned}$ | 3-13+ | Includes annual est. of cannibalism from 1984, o therwise set to 0.2 for all ages in all years |

Tuning data:

| Type | Name | Year range | Age range |
| :--- | :--- | :--- | :--- |
| Tuning fleet1 | Russian com. CPUE, <br> trawl | 1985 - last data year | $9-11$ |
| Tuning fleet2 | Joint Barents Seatrawl <br> survey, february | 1981- last data year | $3-8$ |
| Tuning fleet3 | Joint Barents Sea <br> Acoustic, February+ <br> Lofoten Acoustic <br> survey | 1985-last data year | $3-9$ |
| Tuning fleet4 | Russian bottom trawl <br> survey, November | 1984-last data year | $3-9$ |

$\qquad$

## XSA-settings

| Type of setting | Settings last ye ar | Used this ye ar (why changed) |
| :---: | :---: | :---: |
| Time series weighting | Tapered time weighting power $=3$ over 10 years | The same |
| Recruitment regression model (catchability analys is) | Catchability dependent of stock size for ages $<6$ <br> Regression type $=\mathrm{C}$ <br> Min. 5 points used <br> Survivor estimates <br> shrunk to the population mean for ages $<6$ <br> Catchability independent <br> of age for ages $>=10$ | The same |
| Terminal population estimation | Survivor estimates shrunk towards the mean F of the final 5 years or the 2 oldest ages. <br> S.E. of the mean to which the estimate are shrunk = 1.0 . <br> Minimum standard error for population estimates derived from each fleet $=$ 0.300 . | The same |
| Prior fleet weighting | Prior weighting not applied | The same |

## D. Short-term projection

Model used: Age structured
Software used: MFDP (version 1a) prediction with management option table
Initial stock size: Taken from the XSA for age 4 and older. The recruitment at age 3 for the initial stock and the following 2 years are estimated from survey data and environmental data using the "hybrid model" described in section 1.4.5 in ICES CM 2008/ACOM:01

Natural mortality: average of the three last years or set equal to the values estimated for the terminal year.

Maturity: average of the three last years
$F$ and $M$ before spawning: Set to 0 for all ages in all years
Weight at age in the stock: Predicted by applying (10yr average) annual increments by cohort on last year's observation.

Weight at age in the catch: Predicted by applying (10yr average) annual increments by cohort on last year's observation.

Exploitation pattern: Average of the three last years, scaled by the Fbar (5-10) to the level of the last year, or to the average of the latest 3 years, if there is no clear trend in $F$ and effort.

Intermediate year assumptions: F constraint

Stock recruitment model used: None
Procedures used for splitting projected catches: Not relevant

## E. Medium-term projections

## F. Long-term projections

SPR and YPR calculations

## G. Biological reference points

Introduced 1998: Blim=112000t, Bpa=500000t, Flim=0.7, Fpa=0.42
Adopted in 2003: $\mathrm{Blim}=220000 \mathrm{t}$, $\mathrm{Bpa}=460000 \mathrm{t}$, $\mathrm{Flim}=0.74$, $\mathrm{Fpa}=0.40$

## H. Other issues

Since the 1999 AFWG a new assessment model (Fleksibest-now Gadget) has been used to provide alternative assessments and to describe characteristics of the data for this stock.

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## Annex 4 - Stock Annex - Haddock in Subareas I and II

## Quality Handbook

Haddock in Subareas I and II
Stock specific documentation of standard assessment procedures used by ICES.

Stock:
Working Group:
Date:
Revised by:

Haddock in Subareas I and II (Northeast Arctic)
Arctic Fisheries Working Group
26.04.2009

Alexey Russkikh / Sondre Aaanes

## A. General

## A.1. Stock definition

The North-East Arctic Haddock (Melanogrammus aeglefinus) is distributed in the Barents Sea and adjacent waters, mainly in waters above $2^{\circ}$ Celsius. Tagging carried out in 1953-1964 showed the contemporary area of the Northeast Arctic haddock to embrace the continental shelf of the Barents Sea, adjacent waters and polar front. The main spawning grounds are located along the Norwegian coast and area between $70^{\circ} 30^{\prime}$ and $73^{\circ} \mathrm{N}$ along the continental slope. Larvae extruded are widely drifted over the Barents Sea by warm currents. The 0 -group haddock drifts from the spawning grounds eastwards and northwards and during the international 0-group survey in august it is observed over wide areas in the Barents Sea.Until maturity, haddock are mostly distributed in the southern Barents Sea being their nursery area. Having matured, haddock migrate to the Norwegian Sea.

## A.2. Fis hery

Haddock are harvested throughout a year; in years when the commercial stock is low they are mostly caught as bycatch in cod trawl fishery; when the commercial stock abundance and biomass are high haddock are harvested during their target fishery. On average approximately $25 \%$ of the catch is with conventional gears, mostly longline, which are used almost exclusively by Norway. Part of the longline catches are from a directed fishery.

The fishery is restricted by national quotas. In the Norwegian fishery the quotas are set separately for trawl and other gears. The fishery is also regulated by a minimum landing size, a minimum mesh size in trawls and Danish seine, a maximum by-catch of undersized fish, closure of areas with high density/catches of juveniles and other seasonal and areal restrictions.

In recent years Norway and Russia have accounted for more than $90 \%$ of the landings. Before the introduction of national economic zones in 1977, UK (mainly England) landings made up $10-30 \%$ of the total. Each country fishing for haddock and engaged in the stock assessment provide catch statistic annually. Summary sheets in AFWG Report indicate total yield of haddock by Subareas I, IIa and IIb as well as catch by each country by years. Catch information by fishing gear used by Norway in the haddock fishery is used internally when making estimations at AFWG meeting.

Catch quotas were introduced in the trawl fishery in 1978 and for the fisheries with conventional gears in 1989. Since January 1997 sorting grids have been mandatory for the trawl fisheries in most of the Barents Sea and Svalbard area. Discarding is prohibited. The minimum catching size of haddock is 39 cm in the Russian Economic zone, 44 cm in Norwegian Economic zone; both minimum landing sizes are used by respective fleets in the Svalbard area pursuant to the Svalbard Treaty 1920). The fisheries are controlled by inspections at sea, requirement of reporting to catch control points when entering and leaving the EEZs and by inspections when landing the fish for all fishing vessels. Keeping a detailed fishing log-book on board is mandatory for most vessels, and large parts of the fleet report to the authorities on a daily basis. There is some evidence that the present catch control and reporting systems are not sufficient to prevent discarding and under-reporting of catches.

The historical high catch level of $320,000 \mathrm{t}$ in 1973 divides the time-series into two periods. In the first period, highs were close to $200,000 \mathrm{t}$ around 1956, 1961 and 1968, and lows were between 75,000 and $100,000 \mathrm{t}$ in 1959, 1964 and 1971. The second period showed a steady decline from the peak in 1973 down to the historically low level of $17,300 \mathrm{t}$ in 1984. Afterwards, landings increased to $151,000 \mathrm{t}$ before declining to $26,000 \mathrm{t}$ in 1990. A new increase peaked in 1996 at 174,000 t . The exploitation rate of haddock has been variable.

The highest fishing mortalities for haddock have occurred at intermediate stock levels and show little relationship with the exploitation rate of cod, in spite of haddock being primarily a by-catch in the cod fishery. The exception is the 1990s when morerestrictive quota regulations resulted in a similar pattern in the exploitation rate for both species. It might be expected that good year classes of haddock would attract more directed trawl fishing, but this is not reflected in the fishing mortalities.

Since 2007, estimates of unreported catches (IUU catches) of haddock have been added to reported landings for the years 2002 and onwards. In 2007-2008, two assessments were presented, based on Norwegian and Russian estimates of IUU catches, respectively. The basis for the Norw egian IUU estimates ( $\mathrm{N}-\mathrm{IUU}$ ) is the annual ratio between cod and haddock in the international reported landings from Sub - area I and Division Iib in 2002-2008. These ratios are assumed to be representative of the ratios in the IUU catches. The ratio is applied to the estimated IUU catches of cod in order to get the estimate for haddock. The estimates are similar to those made by the Norwegian Directorate of Fisheries for 2005-2008. The Russian estimates of IUU haddock are obtained by applying the same ratio, but using the Russian estimate of IUU catches of cod in 2002-2007. Both approaches show an increase from 2002 to 2005 followed by a decline. In 2009, the Working Group decided to follow the same procedure used as basis for advice in last year's, and only use the Norwegian IUU.

## A.3. Ecosystem aspects

The composition and distribution of species in the Barents Sea depend considerably on the position of the polar front which separates warm and salty Atlantic waters from colder and fresher waters of arctic origin. Variation in the recruitment of haddock has been associated with the changes in the influx of Atlantic waters to the large areas of the Barents Sea shelf.

In dependence on age and season haddock can vary their diet and act as both predator and plankton-eater or benthos-eater. During spawning migration of capelin (Mallotus villosus) haddock prey on capelin and their eggs on the spawning grounds. When the capelin abundance is low or when their areas do not overlap, haddock can
compensate for lacking capelin with other fish species, i.e. young herring (Clupea harengus) or euphausiids and benthos, which are predominant in the haddock diet throughout a year. Haddock growth rate depends on the population abundance, stock status of main preys and water temperature.
Water temperature at the first and second years of the haddock life cycle is a fairly reliable indicator of year-class strength. If mean annual water temperature in thebottom layer during the first two years of haddock life does not exceed 3.75 C (Kolasection), the probability that strong year-classes will appear is very low even under favourable effect of other factors. Besides, a steep rise or fall of the water temperature shows a marked effect on abundance of year-classes.
Nevertheless, water temperature is not always a decisive factor in the formation of year-class abundance. Strength of year-classes is also determined to a great extent by size and structure of the spawning stock. Under favourable environmental conditions strong year classes are mainly observed in years when the spawning stock is dominated by individuals from older age groups which abundance is at a fairly high level.

Annual consumption of haddock by marine mammals, mostly seals and whales, depends on stock status of capelin as their main prey. In years when the capelin stock is large the importance of haddock in the diet of marine mammals is minimal, while under the capelin stock reduction a considerable increase in consumption by marine mammals of all the rest abundant Gadoid species including haddock is observed (Korzhev and Dolgov, 1999; Bogstad, 2000).

The appearance of haddock strong year classes usually leads to a substantial increase in natural mortality of juveniles as a result of cod predation.

## B. Data

## B. 1 Commercial catch

## Norway

Norwegian commercial catch in tonnes by quarter, area and gear are derived from the sales notes statistics of The Directorate of Fisheries. Data from about 20 sub-areas are aggregated on 6 main areas for the gears gill net, long line, hand line, purse seine, Danish seine, bottom trawl, shrimp trawl and trap. For bottom trawl the quarterly area distribution of the catches is adjusted by logbook data from The Directorate of Fisheries and the total bottom trawl catch by quarter and area is adjusted so that the total annual catch for all gears is the same as the official total catch reported to ICES. No discards are reported or accounted for.

The sampling strategy is to have age and length samples from all major gears in each main area and quarter. The main sampling program is sampling the landings. Additional samples from catches are obtained from the coast guard, from observers and from crew members reporting according to an agreed sampling procedure.
The age distribution and weight at age for the Norwegian catches were estimated using the software based on the method of Hirst et al. (2005). In this method, the three different types of available samples (age and weight samples, age and weight stratified by length groups, and length samples) are modelled simultaneously using a previously developed Bayesian hierarchical model (Hirst et al., 2004). This method replaced the traditional method in 2006, and the time series of Norwegian catch at age (early 80's and onward) was updated based on the modelling approach. The old
method involved allocating unsampled catches to sampled catches based on judgements on "distance criteria's" (in area, time and sometimes gear) and the use of ALK's to fill holes in the sampling frame.

## Russia

Russian commercial catch in tonnes by seasons and area are derived from the Russian Federal Research Institute of Marine Fisheries and Oceanography (VNIRO, Moscow) statistics department. Data from each fishing vessel are aggregated on three ICES sub-Division (I, IIa and IIb). Russian fishery by passive gears was almost stopped by the end of the 1940s. Until late 1990's, relative weight (percentage) of haddock taken by bottom trawls in the total Russian yield exceeded $99 \%$. Only in recent years an upward trend in a proportion of Russian long-line fishery for haddock was observed to be up to $5 \%$ on the average.

The sampling strategy was to conduct mass measurements and collect age samples directly at sea, onboard of both research and commercial vessels to have age and length distributions from each area and season. Data on length distribution of haddock in catches are collected in areas of cod and haddock fishery all the year round by a "standard" fishery trawl (mesh size is $125 / 135 \mathrm{~mm}$ in the Russian Economic zone and Svalbard area and 135 mm in the Norwegian Economic zone) and summarized by three ICES sub-areas (I, IIa and IIb).

Age sampling was carried out by two ways: without any selection (otoliths were taken from any fish caught in one trawl, usually from 100-300 sp.) or using a stratified by length sampling method (i.e. approximately $10-15 \mathrm{sp}$. per each $10-\mathrm{cm}$ length group). The last method has been used since 1988.

All fish taken for age-reading were measured and weighted individually.
Data on length distribution of haddock in catches, as well as age-length keys, are formed for each ICES Subarea, each fishing gear (trawl and longline) for the whole year. Catch at age are reported to ICES AFWG by sub-Division (I, IIa and IIb) for the whole year. In the lack of data by ICES Subareas, information on size-age composition of catches from other areas is used.

## Germany

Catch at age reported to the WG by ICES sub-Division (I, Ila and IIb) according to national sampling. Missing sub-Divisions filled in by use of Russian or Norwegian sampling data.

## Other nations

Total annual catch in tonnes is reported by ICES sub-Divisions or by Russian and Norwegian authorities directly to WG. All catches by other nations are taken by trawl. The age composition from the sampled trawl fleets is therefore applied to the catches by other nations.

The text tablebelow shows which country supplied which kind of data:

|  | KIND OF DATA |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Country | Caton (catch <br> in weight) | Canum (catch <br> at age in <br> numbers) | Weca(weight <br> at age in the <br> catch) | Matprop <br> (proportion <br> mature by <br> age) | Length <br> composition <br> in catch |
| Norway | x | x | x | x | x |
| Russia | x | x | x | x | x |
| Germany | x | x | x | x |  |
| United | x |  |  |  |  |
| King dom | x |  |  |  |  |
| France | x |  |  |  |  |
| Spain | x |  |  |  |  |
| Portugal | x |  |  |  |  |
| Ireland | x |  |  |  |  |
| Greenland | x |  |  |  |  |
| Faroe Islands | x |  |  |  |  |
| Iceland |  |  |  |  |  |
|  |  |  |  |  |  |

The combined catch data were estimated by the SALLOC program (Patterson, 1998). The national data will soon be available in Intercatch, until then the data should be found in the national laboratories and with the stock co-ordinator.

For 1983 and later years mean weight at age in the catch is calculated as the weighted average for the sampled catches. For the earlier period (1946-1982) mean weight at age in catches is set equal to mean weight at age in the catch.
The result files can be found at ICES (sharepoint) and with the stock co-ordinator as ASCII files on the Lowestoft format.

## B.2. Biological

Stock weights used from 1985 and onwards are averages of values derived from Russian surveys in autumn (mostly October-December) and Norwegian surveys in Janu-ary-March the following year. These averages are assumed to give representative values for the beginning of the year. In 2006 the W orking group decided to model the stock weight-at-age data in order to remove some of the sampling variability in the estimates. The weight at age is modelled as follows: Mean length at age is modelled using a von Bertalanffy model with $L_{\infty}$ and $T_{0}$ parameters estimated over the whole time series and a separate K parameter for each year class. Weight at age is estimated from a length-w eight relationship using the smoothed (modelled) length at age. Estimates were produced separately for the Russian autumn survey and thejoint winter survey and were later combined as plain average. For the earlier period (1950-1984) mean weight at age in stock is set equal to mean weight at age in the stock for 1985 and onwards.

In 2006 the Working Group revised the estimates of maturity at age. For the years 1980 onwards the series consists of predicted values using a logistic link function with age and length as explanatory variables from the joint winter survey combined with predicted proportions from the Russian autumn survey:

$$
M a t=\frac{1}{1+e^{(-a *(\text { age-age50\%) }}}
$$

The new series is based on the data from the Russian autumn survey and the joint winter survey. For the period 1950-1979 an average of both data series is used.

For both estimations and predictions the fixed natural mortality of 0.2 is used, and for age 3-6 mortality from predation is applied in addition.

Both the proportion of natural mortality before spawning (Mprop) and the proportion of fishing mortality before spawning (Fprop) are set to 0 . The peak spawning occurs most years in the middle of April.

## B.3. Surveys

## Russia

Russian surveys of cod and haddock in the southern Barents Sea started in the late 1940s as trawl surveys of young demersal fishes. Since 1957 such surveys have been conducted over the whole feeding area including the Bear Island - Spitbergen area (Baranenkova, 1964; Trambachev, 1981), both young and adult haddock have been surveyed simultaneously. In 1984, acoustic methods started to be implemented during surveys of fish stocks (Zaferman, Serebrov, 1984; Lepesevich, Shevelev, 1997; Lepesevich et al., 1999). In 1995 a new acoustic assessment method was applied for the first time, which allowed the differentiation and registration of echo intensities from fish of different length (Shevelev et al., 1998).

Time of survey conducting has reduced from 5-6 months (September-February) in 1946-1981 to 2-2.5 months (October-December) since 1982. The aim of conducting a survey is to investigate both the commercial size haddock as well as the young haddock. The survey covers the main areas where fries settle down as well as the commercial fishery takes place. A total number of more than 400 trawl hauls are conducted during the survey (mainly bottom trawl, a few pelagic trawl).

There are two survey abundance indices at age: 1). absolute numbers (in thousands) computed from the acoustics and 2). trawl indices, calculated as relative numbers per hour trawling. From 1995 onwards there has been a substantial change in the method for calculating acoustic indices. The acoustic survey is therefore presented in 2 tables (Table B4a and B4b) for old and new method of calculating indices.
Ages 1-7 are used in the XSA-tuning.
Norwegian (from 2000 - Joint Norwegian-Russian) winter (February) survey
The survey started in 1981 and covers the ice-free part of the Barents Sea. Both sw ept area estimates from bottom trawl and acoustic estimates are produced. The swept area estimates are used in the tuning for ages 1-8. The survey is described in Jakobsen et al (1997) and Aglen et al. (2002).
Before 2000 this survey was made without participation from Russian vessels, while in the three latest surveys Russian vessels have covered important parts of the Russian zone. The indices for 1997 and 1998, when the Russian EEZ was not covered, have been adjusted as reported previously (Mehl, 1999). The number of fish (age group by age group) in the Russian EEZ in 1997 and 1998 was interpolated assuming a linear development in the proportion found in the Russian EEZ from 1996 to 1999. These estimates were then added to the numbers of fish found in the Norwegian EEZ and the Svalbard area in 1997 and 1998.

It should be noted that the survey conducted in 1993 and later years covered a larger area compared to previous years (Jakobsen et al. 1997). In 1991 and 1992, the number of young cod (particularly 1- and 2-year old fish) was probably underestimated, as cod of these ages were distributed at the edge of the old survey area. Other changes in the survey methodology through time are described by Jakobsen et al. (1997). Note that the change from 35 to 22 mm mesh size in the codend in 1994 is not corrected for in the time series. This mainly affects the age 1 indices.

## B.4. Commercial CPUE

## Russia

No Russian data are used in the stock estimations.

## Norway

Historical time series of observations from onboard Norwegian trawlers were earlier used for tuning of older age groups in VPA. The basis was catch per unit effort (CPUE) in Norwegian statistical areas 03,04 and 05 embracing coastal banks north of the Lofoten, on which approximately $70 \%$ of Norwegian haddock catch fell. However, proportion of haddock taken as by-catch is pretty high and thus it is difficult to estimate their actual catch per unit effort. Since 2002, CPUE indices have not been used in XSA tuning.

## Other data

Not used.

## C. Historical Stock Development

Model used: XSA
Software used: FLR suite and IFAP / Low estoft VPA suite,
Model Options chosen:
Tapered time weighting applied, power $=3$ over 20 years
Catchability independent of stock size for ages $>6$
Catchability independent of age for ages $>=9$
Survivor estimates shrunk towards the mean F of the final 5 years or the 3 oldest ages
S.E. of the mean to which the estimate are shrunk $=0.500$

Minimum standard error for population estimates derived from each fleet $=0.300$
Prior weighting not applied
Input data types and characteristics:

| TYPE | Name | Year range | AGE RANGE | Variable from year to year Yes/No |
| :---: | :---: | :---: | :---: | :---: |
| Caton | Catch in tonnes | 1950 - last data year | 1-11+ | Yes |
| Canum | Catch at age in numbers | 1950 - last data year | 1-11+ | Yes |
| Weca | Weight at age in the commercial catch | 1983 - last data year | 1-11+ | Yes, setequal to west for 19501982 |
| West | Weight at age of the spawning stock at spawning time. | 1950 - last data year | 1-11+ | Yes |
| Mprop | Proportion of natural mortality before spawning | 1950 - last data year | 1-11+ | No - set to 0 for all ages in all years |
| Fprop | Proportion of fishing mortality before spawning | 1950 - last data year | 1-11+ | No - set to 0 for all ages in all years |
| Matprop | Proportion mature at age | 1950 - last data year | 1-11+ | Yes, setequal to average for 19501980 |
| Natmor | Natural mortality | 1950 - last data year | 1-11+ | Includes annual est. of predation by cod from 1984, o therwise set to 0.2 for all ages in all years |

Tuning data:

| TYPE | NAME | Year RANGE | AGE RANGE |
| :--- | :--- | :--- | :--- |
| Tuning fleet1 | Russian bottom trawl <br> survey, October- <br> December | 1983 - last data year | $1-7$ |
| Tuning fleet2 | Joint Barents Sea trawl <br> survey, February | 1982 - last data year | $1-8$ |
| Tuning fleet3 | Joint Barents Sea <br> Acoustic survey, <br> February | 1980 - last data year | $1-7$ |

## D. Short-Term Projection

Model used: Age structured
Software used: R and FLR suite, IFAP prediction with management option table and yield per recruit routines

Initial stock status: is estimated in XSA as abundance of individuals survived in the terminal year for age 3 and older.
Recruitment at age 3 for the start year and the 2 consecutive years is estimated from survey data in RCT3.
Natural mortality is mainly assumed equal to the level estimated for terminal year or to the average for the recent 3 years in dependence on expected cod predation. Method used to determine this parameter and its substantiation are given in the AFWG Reports.
Proportion mature: for current year preliminary actual data presented by Russia are used; for subsequent years - expert estimates by AFWG members. Method used to determine this parameter and its substantiation are given in the AFWG Reports.
$F$ and $M$ prior to spawning are assumed equal to 0 for all ages in all years.
Weight at age in the stock: Method used to determine this parameter and its substantiation are given in the AFWG Reports.

Weight at age in catch: Method used to determine this parameter and its substantiation are given in the AFWG Reports.
Distribution of fishing mortality at age (fishing pattern): For current year it is taken to be at the level of previous year (Fstatus quo) or to be equal to average for the recent 3 years; for subsequent years method used to determine this parameter and its substantiation are given in the AFWG Reports.
$F$ and Mbefore spawning: Set to 0 for all ages in all years
Stock recruitment model used: None
Procedures used for splitting projected catches: Not relevant

## E. Medium-Term Projections

Time lag: 4 years
Software used: R and FLR.

Initial stock status, natural mortality, proportion mature, proportion of F and M prior to spawning, mean weight at age in stock and in catch, exploitation pattern, predicted $F$ in intermediate year: the same as in the short-term prediction.

Stock recruitment model is not used.
Uncertainty models used: See AFWG 2007.

## F. Long-Term Projections

Spawning stock biomass per recruit (SPR) and yield per recruit (YPR) are estimated annually.

## G. Biological Reference Points

Introduced 1998: Blim=50000t, Bpa=80000t, Flim=0.49, Fpa=0.35

## H. Other Issues

## Harvest control rule

The harvest control rule (HCR) was evaluated by ICES in 2007 (AFWG 2007) and found to be in agreement with the precautionary approach. The agreed HCR for haddock is as follows (Protocol of the 36th Session of The Joint Norwegian Russian Fishery Commission, 10 October 2007):

- TAC for the next year will be set at level corresponding to $F_{p a}$
- The TAC should not be changed by more than +/- $25 \%$ compared with the previous year TAC.
- If the spawning stock falls below $B_{p a}$, the procedure for establishing TAC should be based on a fishing mortality that is linearly reduced from $F_{p a}$ at $B_{p a}$ to $F=0$ at $S S B$ equal to zero. At SSB-levels below $B_{p a}$ in any of the operational years (current year and a year ahead) there should be no limitations on the year-to-year variations in TAC.


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## Annex 5 - Stock Annex - Northeast Arctic Saithe

## Quality Handbook

Annex: Saithe in Subareas I and II
Stock specific documentation of standard assessment procedures used by ICES.

Stock: $\quad$ Saithe in Subareas I and II (Northeast Arctic)
Working Group: Arctic Fisheries Working Group
Date:
28.04.2010

Revised by:
Sigbjørn Mehl / Åge Fotland

## A. General

## A.1. Stock definition

The Northeast Arctic saithe is mainly distributed along the coast of Norway from the Kola Peninsula in northeast and south to Stad at $62^{\circ} \mathrm{N}$ (Figure 1). The 0 -group saithe drifts from the spawning grounds to inshore waters. 2-4 years old the saithe gradually moves to deeper waters, and at age $3-6$ it is found at typical saithe grounds. It starts to mature at age 5-7 and in early winter a migration towards the spawning grounds further out and south starts.

The stock boundary $62^{\circ} \mathrm{N}$ is more for management purposes than a biological basis for stock separation. Tagging experiments show a regular annual migration of mature fish from the North-Norwegian coast to the spawning areas off the west coast of Norway and also to a lesser extent to the northern North Sea (ICES 1965). There is also a substantial migration of immature saithe to the North Sea from the Norwegian coast betw een $62^{\circ}$ and $66^{\circ} \mathrm{N}$ (Jakobsen 1981). In some years there are also examples of mass migration from northern Norway to Iceland and to a lesser extent to the Faroe Islands (Jakobsen 1987). 0-group saithe, on the other side, drifts from the northern North Sea to the coast of Norway north of $62^{\circ} \mathrm{N}$.

## A.2. Fis hery

Norway accounts for more than $90 \%$ of the landings. Over the last ten years about $40 \%$ of the Norw egian catch originates from bottom trawl, $25 \%$ from purse seine, $20 \%$ from gill net and $15 \%$ from other conventional gears (long line, Danish sine and hand line). The gill net fishery is most intense during winter, purse seine in the summer months while the trawl fishery takes place more evenly all year around. Landings of saithe were highest in 1970-1976 with an average of $239,000 \mathrm{t}$ and a maximum of $265,000 \mathrm{t}$ in 1974 (Figure 2). Catches declined sharply after 1976 to about $160,000 \mathrm{t}$ in the years 1978-1984. This was partly caused by the introduction of national economic zones in 1977. The stock was accepted as exclusively Norwegian and quota restrictions were put on fishing by other countries while the Norwegian fishery for some years remained unrestricted. Another decline followed and from 1985 to 1991 the landings ranged from 67,000 to $123,000 \mathrm{t}$. An increasing trend was seen after 1990 to $171,000 \mathrm{t}$ in 1996 , followed by a new decline to $136,000 \mathrm{t}$ in 2000. Since then the annual landings have increased gradually to $212,000 \mathrm{t}$ in 2006, followed by a decline to

199000 t in 2007, 183000 t in 2008 and 161000 t in 2009. Quotas can be transferred between gears if the quota allocated to one of the gears will not be taken. The target set for the total landings has generally been consistent with the scientific recommendations.


Figure 1. NEA saithe. Distribution of larvae, juveniles, adult spawning areas and the main migration patterns by (a) first quarter, (b) second quarter, (c) third quarter, and (d) fourth quarter.

The number of vessels taking part in the purse seine fishery has varied between 110 and 429 since 1977, with the highest participation in the first part of the period. There have been some variations from year to year, and many of the vessels that have taken
part in the fishery the last decade have accounted for only a small fraction of the purse seine catches. The annual effort in the Norwegian trawl fishery has varied between 12000 and 77000 hours, with the highest effort from 1989 to 1995. Like in the purse seine fishery there have been rather large changes from year to year.


Figure 2. NEA saithe landings 1960-2009. Red part of bars shows the Norwe gian landings.

1 March 1999 the minimum landing size was increased from $35-40 \mathrm{~cm}$ to 45 cm for trawl and conventional gears, and to 42 cm (north of Lofoten) and 40 cm (between $62^{\circ}$ N and Lofoten) for purse seine, with an exception for the first 3000 t purse seine catch between $62^{\circ} \mathrm{N}$ and $66^{\circ} 33^{\prime} 30 \mathrm{~N}$, where the minimum landing size still is 35 cm .

## A.3. Ecosystem aspects

The recruitment of saithe may suffer in years with reduced inflow of Atlantic water (Jakobsen 1986).

## B. Data

## B.1. Commercial catch

Norwegian commercial catch in tonnes by quarter, area and gear are derived from the sales notes statistics of The Directorate of Fisheries. Data from about 20 sub areas are aggregated on 6 main areas for the gears gill net, long line, hand line, purse seine, Danish seine, bottom trawl, shrimp trawl and trap. For bottom trawl the quarterly area distribution of the catches is adjusted by logbook data from The Directorate of Fisheries and the total bottom trawl catch by quarter and area is adjusted so that the total annual catch for all gears is the same as the official total catch reported to ICES. No discards are reported or accounted for, but there are several reports of discards. In later years there are also reports of misreporting, saithe is landed as cod in a period with decreasing quotas and availability of cod and good availability of saithe.

The sampling strategy is to have age-length samples from all major gears in each area and quarter. There are at present no defined criteria on how to allocate samples of catch numbers, mean length and mean weight at age to unsampled catches, but the
following general process has been applied: First look for samples from a neighbouring area if the fishery extends to this area in the same quarter. If there are no samples available in neighbouring areas, search for samples from other gears with the most similar selectivity in the same area or in neighbouring areas. The last option is to search in neighbouring quarters, first from the same gear in the same area, and then from neighbouring areas and similar gears. For some gears, areas and quarters length samples taken by the coast guard are applied and combined with an ALK from a neighbouring area, gear or quarter. ALKs from research surveys (shrimp trawl) are also used to fill holes. The alternative method applied for cod and haddock (ECA, Hirst et al. 2004, 2005) produce unrealistic high weights at age compared to the method presently applied for NEA saithe (ICES 2007/ACFM:16).

Constant weight at age values is used for the period 1960 - 1979. For subsequent years, Norw egian weights at age in the catch are estimated from length at age by the formula:

$$
\text { Weight }(\mathrm{kg})=\left(1^{3 *} 5.0+1^{2} * 37.5+1^{*} 123.75+153.125\right)^{*} 0.0000017,
$$

Where

$$
\mathrm{l}=\text { length in } \mathrm{cm} .
$$

Norway has on average accounted for about $95 \%$ of the saithe landings. Data on catch in tonnes from other countries are either taken from ICES official statistics (by ICES area) or from reports to Norwegian authorities. A few countries also supply some additional data. The text table below shows which countries supply which kind of data:

|  | KIND OF DATA |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Country | Caton (catch in weight) | Canum (catch at age in numbers) | Weca (weight at age in the catch) | Matprop (proportion mature by age) | Leng th composition in catch |
| Norway | x | x | x | X | X |
| Russia | x | x | x |  | x |
| Germany | x | x | x |  |  |
| United king dom | x |  |  |  |  |
| France | x |  |  |  |  |
| Spain ${ }^{1}$ | X |  |  |  |  |
| Portugal | x |  |  |  |  |
| Poland | x |  |  |  |  |
| Greenland ${ }^{1}$ | x |  |  |  |  |
| Faroe Islands ${ }^{1}$ | x |  |  |  |  |
| Iceland ${ }^{1}$ | X |  |  |  |  |

${ }^{1}$ As reported to Norwegian authorities

The Norwegian, Russian and German input files are Excel spreadsheet files. Russian input data earlier than 2002 are supplied on paper and later punched into Excel spreadsheet files before aggregation to international data. The data should be found in the national laboratories and with the Norwegian stock co-ordinator.
The national data have been aggregated to international data on Excel spreadsheet files. Age composition data are normally available from Norway, Russia (some areas) and Germany (Division IIA). In some areas Russian length composition has been ap-
plied on the Russian landings together with an age-length-key (ALK) and weight at age data from the Norwegian trawl landings. Catches from the other countries were assumed to have the same age composition and weight at age as the Norw egian trawl landings. In some years the final German and Russian numbers at age have been adjusted to remove SOP discrepancies before aggregation to international data. The Excel spreadsheet files used for age distribution, adjustments and aggregations can be found with the Norwegian stock co-ordinator. Since 2007 the national data have also been uploaded to the ICES InterCatch database.
The result files (FAD data) can be found with the stock co-ordinator and at ICES as ASCII files on the Lowestoft format under w:lacomlafwg\yearlStock|sai_arct.

## B.2. Biological

Weight at age in the stock is assumed to be the same as weight at age in the catch.
A fixed natural mortality of 0.2 is used both in the assessment and the forecast.
Both the proportion of natural mortality before spawning (Mprop) and the proportion of fishing mortality before spawning (Fprop) are set to 0 .

Regarding the proportion mature at age, until AFWG 1995 knife-edge maturity at age 6 was used for this stock. In the 1996-2004 assessments, an ogive based on analyses of spawning rings in otholiths for the period 1973-1994 was applied for all years. The analysis showed a lower maturation in the last part of the period, and some extra weight was given to this part when an average ogive was calculated. In 2005 a large number of otholiths with missing information on spawning rings were re-read, and new analyses were done for the period 1985-2004. The maturity at age had decreased somewhat in the last part of that period, and the 2005 WG decided to use a 3-year running average, reference year being the middle of the 3-year period, for the years from 1985 and onwards (2-year average for the first and last year) (ICES 2005). The ogives used until AFWG 1995 and in 1996-2004 assessments are presented in the text table below.

| AGE GROUP | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1}+$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Until 1995 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1996-2004 | 0 | 0 | 0.01 | 0.55 | 0.85 | 0.98 | 1 | 1 | 1 | 1 |

## B.3. Surveys

In 1985-2002 a Norwegian acoustic survey specially designed for saithe was been conducted annually in October-November (Nedreaas 1997). The survey covers the near coastal banks from the Varangerfjord close to the Russian border and southwards to Stad at $62^{\circ} \mathrm{N}$ (Figure 3). The whole area has been covered since 1992, and the major parts since 1988. The aim of conducting an acoustic survey targeting Northeast Arctic saithe has been to support the stock assessment with fisheryindependent data of the abundance of the youngest saithe. The survey mainly covers the grounds where the trawl fishery takes place, normally dominated by 3-5(6) year old fish. 2-year-old saithe, mainly inhabiting the fjords and more coastal areas, are also represented in the survey, although highly variably from year to year. In 1997 and 1998 there was a large increase in the abundance of age 5 and older saithe, confirming reports from the fishery. In 1999 the abundance of these age groups decreased somewhat, but was still at a high level compared to the years before 1997 (Mehl 2000). Abundance indices for ages 2-5 were used for tuning from 1988 on-
wards, but including older ages as a 6+ group in the tuning series improved the scaled weights a little and at the 2000 WG meeting it was decided to apply the extended series in the assessment. The results from the survey in autumn 2000 showed a further decrease in the abundance of age 5 and older saithe (Korsbrekke and Mehl 2000). It is not known how well the survey covers the oldest age groups from year to year, but at least for precautionary reasons the $6+$ group was kept in the tuning series. Before the 2005 WG the 6+ group from the Norwegian acoustic survey was split into individual age groups $6-9$ by rerunning the original acoustic abundance estimates. However, this was only possible to do for the years back to 1994. Based on further analysis during the 2005 benchmark assessment, indices for ages 3-7 was used for tuning in the 2005 and later assessments.


Figure 3. NEA saithe. Distribution of total saithe echo density in the acoustic survey autumn 1998.

In 1995-2002 a Norwegian acoustic survey for coastal cod was conducted along the coast and in the fjords from Varanger to Stad in September, just prior to the saithe survey described above. This survey covers coastal areas not included in the regular saithe survey. Because saithe is also acoustically registered, this survey provides supplementary information, especially about 2- and 3-year-old saithe that have not yet migrated out to the banks. At the WG meeting in 2000 analyses were done on com-
bining these indices with indices from the regular saithe survey in the tuning series, but it did not influence the assessment much. The WG therefore decided, for the time being, to apply only indices from the longer time series of the regular saithe survey in the assessment.


Figure 4. Standard transects in new combine d saithe and coastal surve y.

In autumn 2003 the saithe- and coastal cod surveys were combined. A new survey was designed, with new stratification and smaller strata based on depth and fish distribution in recent years, and with new and more regular transects (Figure 4). The new course lines had already been partly introduced in the saithe survey in 2001 and 2002. At the 2010 benchmark assessment two alternative survey index series was tested, one for 2001-2008 representing the traditional saithe survey area with new course lines and stratification, and one for 2003-2008 representing the combined saithe and coastal cod survey areas. The new tuning series gave lower and more stable S. E. Log q residuals than the tuning series presently used. However, the retrospective trend was still poor and the estimates of F and SSB in the last assessment year were far away from any other analysis. The new series are probably still too short to be used for tuning of the NEA saithe XSA. Until a longer time series based on the new survey design is established, indices from the whole survey time series, rep-
resenting the traditional saithe survey area only, will be applied for tuning. The estimation of these abundance indices is done very much in the same way for the whole time series and the results for later years should be comparable with earlier years.

## B.4. Commercial CPUE

Two CPUE data series have been used, one from the Norwegian purse seine fishery and one from the Norw egian trawl fishery.

Until 1999 indices of fishing effort in the purse seine fishery were based on the number of vessels of 20-24.9 m length and the effort (number of vessels) of this length category was raised by the catches to represent the total purse seine effort. However, the number of vessels taking part in the fishery almost doubled from 1997 to 1998, but due to regulations the catches were almost the same as in 1997. In such a situation the total number of vessels participating in a fishery is clearly not a good measure of effort. Examination of the data showed that many of the vessels that have taken part in the fishery the last decade have accounted for only a small fraction of the purse seine catches, and these also included most of the vessels that tend not to be involved on a regular basis. Roughly half of the vessels have caught less than 100 tonnes per year, and the sum of these catches represents only about $5-10 \%$ of the total purse seine catch. Therefore the number of vessels catching more than 100 tonnes annually seems to be a more representative and more consistent measure of effort in the purse seine fishery. These numbers are raised to the total purse seine catch. The new effort series showed a smaller decrease in later years than the old one and in the XSA runs it gets higher scaled weights. The 2000 WG meeting therefore decided to use the new CPUE data series in the assessment.

The quality and performance of the purse seine tuning fleet has been discussed several times in the WG. The effort, measured as number of vessels participating, has been highly variable from year to year. This was partly taken care of by only including vessels with total catch > 100 tonnes. How ever, with a restricting and changing TAC and transfer of quota, the CPUE may change much from year to year without really reflecting trends in the saithe abundance. This is also reflected in the tuning diagnostics of exploratory runs. There are rather large and variable $\log \mathrm{q}$ residuals and large S.E. $\log \mathrm{q}$ for all age groups except age 4 , which often is the dominant age group in the purse seine landings. But even for age 4 the S.E. $\log \mathrm{q}$ is higher than in the Norwegian trawl CPUE and acoustic survey indices single fleet tunings. There are strong year effects, and in the combined tuning the purse seine series get low scaled weights. Mainly based on this the 2005 WG decided to not include the purse seine tuning fleet in the analysis (ICES 2005). In later years with lower availability of young saithe the TAC has been less restricting, and at the 2010 benchmark assessment exploratory runs were done with updated purse seine tuning series. The purse seine tuning series showed the higher S.E_Log q residuals and lower scaled weights than the other tuning series and did not perform any better than in previous analysis, and were not reintroduce as a tuning series in the assessment.

Catch and effort data for Norwegian trawlers were until 2000 taken from hauls where the effort almost certainly had been directed towards saithe, i.e., days with more than $50 \%$ saithe and only on trips with more than $50 \%$ saithe in the catch. The effort estimated for the directed fishery was raised by the catches to give the total effort of Norwegian trawlers. From 1997 to 1998 the effort increased by more than $50 \%$, but due to regulations the catches were slightly low er in 1998 and the CPUE decreased by almost $40 \%$ from 1997 to 1998 and stayed low in 1999. This may at least partly be ex-
plained by change in fishing strategies in a period with increasing problems with bycatch of saithe in the declining cod fishery due to good availability of saithe. In 2001 new CPUE indices by age were estimated based on the logbook database of the Directorate of Fisheries, which has a daily resolution (Salthaug and Godø 2000). After some initial analyses it was decided to only include data from vessels larger than the median length since they showed the least noisy trends. One single CPUE observation from a given vessel is the total catch per day divided by the duration of all the trawl hauls that day. To increase the number of observations during a time period with decreasing directed saithe fishery, all days with $20 \%$ or more saithe were included. The effort (hours trawling) for each CPUE observation was standardised or calibrated to a standard vessel. Until 2002, first averaging all CPUE observations for each month, and then averaging over the year a yearly index was calculated. The CPUE indices were divided on age groups by quarterly weight, length and age data from the trawl fishery. From 2003, first averaging all CPUE observations for each quarter, and then averaging over the year a yearly index was calculated. The CPUE indices were finally divided on age groups by yearly catch in numbers and weight at age data from the trawl fishery. The new approach was less influenced by short periods with poor data, while it still evens out seasonal variations.

There was an increase in the total CPUE from 1999 to 2003, when it reached the highest level in the time series going back to 1980. In 2004 the total CPUE was almost exactly the same as in 2003, while there was about a $30 \%$ increase from 2004 to 2005. This was caused by an increase in the quarter one CPUE. This increase started already in 2003, but was most pronounced in 2005. The increase may be explained by increased availability and catchability of saithe in spawning areas of Norwegian spring spawning herring, where the saithe feeds on herring during quarter one. A similar increase was not seen in the other areas and quarters. AT the 2005 benchmark assessment an annual CPUE series was calculated without quarter one data. This CPUE series showed much less variations over the last four years, and the WG decided to use a CPUE time series averaged over quarters 2-4 for tuning (ICES 2005). Due to rather large negative log q residuals in the first part of the new time series, it was shortened to only cover the period after 1993. Based on exploratory runs done at the 2005 benchmark assessment the age span was set to 4-8.

The estimates of total CPUE increased considerably both in 2007 and 2008. The survey (Aglen et al. 2009) shows a higher proportion of saithe in the southern half of the distribution area in the last years, and logbook data show that the trawl catches included in the CPUE calculations also have become gradually more southerly distributed, i.e. the trawlers follow saithe aggregations that may have become extra available in 2007 and 2008. The biological samples used for dividing total CPUE on age groups are, however, from the whole saithe fishery and therefore include age groups that are not numerous in these aggregations. Based on this and the decline in survey indices in the same years and additional analysis, the WG decided to exclude the 2007 and 2008 CPUE data in the final assessment (ICES 2008, ICES 2009a).

Further analysis and exploratory runs were presents at the 2010 benchmark assessment. Six different options were tested, included a proposal from the industry. The CPUE index based upon 7 vessels proposed by the industry could implement new bias or noise due to lack of quarterly indices and index values out of range. To take account of a time period (2000-2008) with increasing directed saithe fishery (Figure $2 \mathrm{~b})$, all days with $80 \%$ or more saithe are excluded in some runs. Of the two options A) leaving out quarter 1 in the averaging and use all catches with $>20 \%$ saithe for the rest of year (as in the current index) or B) leaving out days with $>20 \%$ but $<80 \%$
saithe and including quarter 1 in the averaging, option $B$ was chosen because it gave somewhat better diagnostics in the XSA runs and is more consistent regarding how data is selected and direct fishery is treated in the rest of the year. The increase in CPUE at the end of the time period was much less for this option and all data years were included in the analysis.


Figure 5a Distribution of small and large trawl catches of NEA saithe (in percent) 1994-1999.


Figure 5a Distribution of small and large trawl catches of NEA saithe (in percent) 2000-2008.

## B.5. Other relevant data

None.

## C. Historical Stock Development

Until the 2005 assessment age 2 was applied as recruitment age in the XSA runs, projections and calculations of reference points. Since the mid 1990's there has been almost no catch of 2 year olds and this age group should in theory be fully protected by the new minimum landing size. 2-year-old saithe, mainly inhabiting the fords and more coastal areas, are represented in the survey, but highly variable from year to year. The saithe is normally not fully recruited to the survey before at age 3 and in some years at age 4. It is therefore difficult to estimate good recruitment indices, even at age 2. This especially effects the projections. Retrospective XSA analyses showed that applying age 3 as recruitment age implies that one may include more years in the last part of the recruitment time series. The 2005 WG therefore decided to apply age 3 as recruitment age.
Since about year 2000 the number of old (11+) fish in the catch matrix has been gradually increasing until 2004 and then decreased somewhat, but is still on a high level compared to the years before 2000. VPA based assessment models fitted to data sets with significant numbers in the oldest age and plus group, are extremely sensitive to the method by which fishing mortality at the oldest age is estimated, due to relatively poor VPA convergence at the oldest ages (see ICES 2002, Annex 7). At the 2010 benchmark assessment (WKROUND 2010) the catch matrix was extended to 15+ to avoid some of the potentially plus group problems. At WKROUND this was only possible to do back to 1989. Exploratory XSA runs showed much better retrospective patterns and lower SSB levels and higher F levels at the end of the time period. Prior to AFWG 2010 the whole time series of both catch, weight and maturity at age was extended.

Analysis of the tuning series indicated that there had been a shift in catchability around year 2002 (Figure 6). The survey was redesigned in 2003, and the fishery to a larger degree targeted older ages. Permanent breaks were made in both tuning series in 2002. This allows the XSA freedom to estimate different qs. Exploratory XSA runs showed improvement of retrospective patterns and diagnostics, and some year effects were no more apparent. Additional exploratory runs with reduced shrinkage were done to better allow the model to fit population number to the tuning series. Detailed XSA diagnostics indicated that both tuning indices were relative good in estimating year class strength at different ages. Therefore lowering the shrinkage, allowing the commercial CPUE and survey to determine more of the year classes seemed appropriate (ICES 2009b). The proposed shrinkage of 1.5 lowered the weight of the shrinkage to less than $4 \%$ for all ages. The use of a 20 year tricubic taper against a no-taper was also investigated. Although diagnostics did not substantially improve, it was decided that there were no benefits in keeping the tricubic taper as the splitting up of the tuning series already had a similar impact on the assessment as the 20 year taper and improved substantially the assessment.

The recommendation from WKROUND 2010 therefore was to run the XSA with a 15+ catch matrix, tuning time series broken in 2002, reduced shrinkage (S.E. of the mean to which the estimate are shrunk increased from 0.5 to 1.5) and no tapered time weighting. The new model options are shown below.


Figure 6 Catchability (index/N) at age in the Norwegian acoustic survey (upper panel) and in the Norwegian trawl CPUE series (lower panel).

Until the 2005 assessment age group 3-6 was the reference age group for Fbar and has been applied in the projections and calculations of fishing mortality reference points. Before the mid 1990's 3 year old fish made up a significant part of the landings, and age group 3-6 contributed about $80 \%$. Since the mid 1990's there has been a marked reduction in the landings of 3 year olds, and age group 4-7 contributes more than age group 3-6. This is partly related to transference of quota from purse seine to conventional gears and partly to better price for larger saithe. In 1999 the minimum landing size was increased, and most of the 3-year-old fish will be below this size the whole year. The 2005 WG therefore decided to apply age group 4-7 as reference age group for Fbar. The fishing mortality PA-reference points therefore were re-calculated.

Due to the increased number of old fish in the catch matrix the 2010 benchmark assessment also investigated the age span for Fbar. Age groups 4-7 still make up most of the landings, and there are more noisy data in older age groups. Therefore it was decided keep Fbar as current.

Model used: XSA
Software used: Lowestoft VPA suite. In AFWG 2009 exploratory assessment runs were conducted in FLR version 2.8.1.

Model Options chosen:
No tapered time weighting applied
Catchability independent of stock size for all ages

Catchability independent of age for ages $>=8$
Survivor estimates shrunk towards the mean F of the final 5 years or the 5 oldest ages S.E. of the mean to which the estimate are shrunk $=1.500$

Minimum standard error for population estimates derived from each fleet $=0.300$
Prior weighting not applied

Input data types and characteristics:

| TYPE | NAME | YEAR RANGE | AGE RANGE | VARIABLE FROM <br> YEAR TO YEAR <br> YES/No |
| :--- | :--- | :--- | :--- | :--- |
| Caton | Catch in tonnes | $1960-$ last data <br> year | $3-15+$ | Yes |
| Canum | Catch at age in <br> numbers | $1960-$ last data <br> year | $3-15+$ | Yes |
| Weca | Weight at age in <br> the commercial <br> catch | $1960-$ last data <br> year | $3-15+$ | Yes/No -constant <br> at age from 1960 - <br> 1979 |
| West | Weight at age of <br> the spawning <br> stock at spawning <br> time. | $1960-$ last data <br> year | $3-15+$ | Yes/No - assumed <br> to be the same as <br> weight at age in <br> the catch |
| Mprop | Proportion of <br> natural mortality <br> before spawning | $1960-$ last data <br> year | $3-15+$ | No - set to 0 for <br> all ages in all <br> years |
| Fprop | Proportion of <br> fishing mortality <br> before spawning | $1960-$ last data <br> year | $3-15+$ | No - set to 0 for <br> all ages in all <br> years |
| Matprop | Proportion mature <br> at age | $1960-$ last data <br> year | $3-15+$ | Yes/No -constant <br> ogive 1960-1984, <br> three year <br> running average <br> since 1985 |
| Natmor | Natural mortality | 1960 - last data |  |  |
| year | $3-15+$ | No - set to 0.2 for <br> all ages in all <br> years |  |  |

Tuning data:

| TYPE | NAME | Year RANGE | AGE RANGE |
| :--- | :--- | :--- | :--- |
| Tuning fleet 11 | Nor trawl quarter 1-4 | $1994-2001$ | $4-8$ |
| Tuning fleet 12 | Nor trawl quarter 1-4 | 2002 - last data year | $4-8$ |
| Tuning fleet 13 | Norway ac survey | 1994 - 2001 | $3-7$ |
| Tuning fleet 14 | Norway ac survey | 2002 - last data year | $3-7$ |

For analysis of alternative procedures seeWG reports from AFWG 1997-2009.

Model used: Age structured
Software used: MFDP prediction with management option table and yield per recruit routines, MFYPR.

Initial stock size. Taken from the XSA for age 5 and older. The recruitment at age 3 in the last data year is estimated using the long-term geometric mean, and numbers at age 4 in the intermediate year is calculated applying a natural mortality of 0.2 and the F value estimated by XSA, (advised by RG in 2004).

From AFWG 2009 the numbers at age 4 in the intermediate year is calculated applying a natural mortality of 0.2 and the $F$ value estimated by standard Pope's equation for calculation of this $y-c$ at age 4 , i.e. $N(4)=\left[N(3)^{*} \exp (-\mathrm{M} / 2)-\mathrm{C}(3)\right]{ }^{*} \exp (-\mathrm{M} / 2)$, (advised by RG in 2009).

Natural mortality:Set to 0.2 for all ages in all years
Maturity: Constant ogive 1960-1984, three year running aver age since 1985, reference year being the middle
$F$ and $M$ before spawning: Set to 0 for all ages in all years
Weight at age in the stock: Assumed to be the same as weight at age in the catch
Weight at age in the catch: For weight at age in stock and catch the average of the last three years in the VPA is normally used.

Exploitation pattern: The average of the last three years for ages 3-10, and a constant value for age 11 to $15+$ calculated as the average of ages $11-13$ over the last three years.

Selection pattern for yield per recruit: The average selection pattern from the last three years (2006-2008) of the assessment was used.

Intermediate year assumptions: TAC constraint, scaled to a TAC value. If using Sq F for the intermediate year, exploitation patterns described above should be used if there is no trend in F. If a trend in F is observed, the exploitation pattern should be scaled by the Fbar (4-7) to the level of the last year.

Stock recruitment model used: None, the long-term geometric mean recruitment at age 3 is used
Procedures used for splitting projected catches: Not relevant

## E. Medium-Term Projections

The issue was not addressed during the 2010 benchmark and no projections were made. Settings previously used are listed below.

Model used: Age structured
Software used: MFDP single option prediction
Initial stock size: Same as in the short-term projections.

Natural mortality: Set to 0.2 for all ages in all years
Maturity: Same as in the short-term projections.
F and Mbefore spawning: Set to 0 for all ages in all years
Weight at age in the stock: Assumed to be the same as weight at age in the catch
Weight at age in the catch:Same as in the short-term projections.
Exploitation pattern: Same as in the short-term projections.
Intermediate year assumptions: F-factor from the management option table corresponding to the TAC

Stock recruitment model used: None, the long-term geometric mean recruitment at age 3 is used

Uncertainty models used: @RISK for Excel, Latin Hyper cubed, 5000 replications, fixed random number generator

- Initial stock size: Lognormal distribution, LOGNORM (mean, standard deviation), with mean as in the short-term projections and standard deviation calculated by multiplying the mean by the external standard error from the XSA diagnostics (except for age 3, see recruitment below)
- Natural mortality: Set to 0.2 for all ages in all years
- Maturity: Constant ogive 1960-1984, three year running average since 1985
- F and Mbefore spawning: Set to 0 for all ages in all years
- Weight at age in the stock: Assumed to be the same as weight at age in the catch
- Weight at age in the catch: Average weight of the three last years
- Exploitation pattern: Average of the three last years, scaled by the Fbar (4-7) to the level of the last year if there is a trend
- Intermediate year assumptions: F-factor from the management option table corresponding to the TAC
- Stock recruitment model used: specified as a PERT distribution (as special form of the beta distribution) with a minimum and maximum value as specified. The shape parameter is calculated from the defined most likely value.

RiskPertAlt(arg1type, arg1value, arg2type,arg2value, arg3type,arg3value). Specifies a PERT distribution with three arguments of the type arg1type to arg3type. These arguments can be either a percentile between 0 and 1 or "min", " $m$. likely" or "max".

Examples: RiskPertAlt(2\%; min; 50\%; geomean; 98\%; max) specifies a PERT distribution with a minimum of min and a most likely value of geomean and a 98th percentile of max.

## F. Long-Term Projections

The issue was not addressed during the 2010 benchmark and no projections were made.

## G. Biological Reference Points

Due to the change of Fbar from 3-6 to 4-7 and age at recruitment from 2 to 3, the lim and pa reference points were re-estimated at the 2005 WG . The $\lim$ reference points were estimated according to the new methodology outlined in ICES CM 2003/ACFM:15. Saithe retrospective XSA-analyses show that in later years there have been an overestimation of F and underestimation of SSB in the assessment year. The trend may have been the opposite in earlier years, but the length of the tuning series do not allow for long enough retrospective analysis to verify this. The new methodology (ICES CM 2003/ACFM:15) does not give any advise on how to deal with such situations. The pa reference point estimation was therefore based on the old procedure, applying the "magic formula" $\mathbf{B}_{\mathrm{pa}}=\mathbf{B}_{\lim } \exp \left(1.645^{*} \sigma\right)$ and $\mathbf{F}_{\mathrm{pa}}=\mathbf{F}_{\text {lim }}{ }^{*} \exp \left(-1.645^{*} \sigma\right)$, where $\sigma$ is a measure of the uncertainty of $F$ estimates (ICES CM 1998/ACFM:10). For NEA saithe a value of 0.3 was applied in both estimates.
In 2010 the age span was expanded from 11+ to 15+ and important XSA parameter settings were changed (ICES CM 2010/ACOM:36). This resulted in changes in estimated fishing mortality, spawning stock biomass and recruitment, especially in the last part of the time series. Therefore the $\lim$ and pa reference points were reestimated at the 2010 WG . The results of the segmented regression were not very much different from the previous analyses. The HCR is based on the PA reference points, and if new ones are introduced, the HCR would have to be evaluated again. Due to lack of time to do this during the WG and the transition to MSY based reference points (see Section 0), it was decided to not change the existing LIM and PA reference points. The estimations done at the present WG are, however, presented below.

## Biomass reference points

In 1994 the WG proposed a MBAL of $150,000 \mathrm{t}$, based on the frequent occurrence of poor year classes below this level of SSB. The new maturity ogive introduced in 1995 gave somewhat higher historical SSB estimates. 150,000 t was considered to represent a less restrictive MBAL and 170,000 $t$ was found to correspond better with the arguments used in 1994 (ICES 1996/Assess: 4). The Study Group on the Precautionary Approach to Fisheries Management (SGPAFM, ICES 1998/ACFM: 10) also found this to be a suitable level for $\mathrm{B}_{\text {pa. }}$. However, based on a visual examination of the stockrecruitment plot ACFMlater reduced the $B_{\text {pa }}$ to 150,000 t (ICES 1998b).

At the 2005 WG parameter values, including the change-point ( $\mathbf{S}^{*}=\mathbf{B}$ lim), slope in the origin $(\hat{\alpha})$ and recruitment plateau $\left(\mathbf{R}^{*}\right)$, were computed using segmented regression on the 1960-2000 time series of SSB-recruitment pairs. The values are presented in the text table below.

| From algorithm in Julious (2001) |  |  |
| :--- | :--- | :--- |
| $\mathrm{S}^{*}$ | $\hat{\alpha}$ | $\mathrm{R}^{*}$ |
| 136378 | 1.27 | 173200 |

Applying the "magic formula" $\mathbf{B}_{\mathrm{pa}}=\mathbf{B}_{\lim } \exp \left(1.645^{*} \sigma\right)$, gives a $\mathbf{B}_{\mathrm{pa}}$ of $223,392 \mathrm{t}$, rounded to $220,000 \mathrm{t}$.

At the 2010 WG this procedure was repeated based on the results of the new assessment settings, using segmented regression on the 1960-2005 time series of the new SSB-recruitment pairs. The new values were:

| From algorithm in Julious (2001) |  |  |
| :--- | :--- | :--- |
| $\mathrm{S}^{*}$ | $\hat{\alpha}$ | $\mathrm{R}^{*}$ |
| 118542 | 1.48 | 175485 |

Applying the "magic formula" $\mathbf{B}_{\mathrm{pa}}=\mathbf{B} \lim \exp \left(1.645^{*} \sigma\right)$, gives a $\mathbf{B}_{\mathrm{pa}}$ of $194,176 \mathrm{t}$. However, as explained above, the existing values of $B_{\lim }=136,000 t$ and $B_{p a}=220,000 \mathrm{t}$ will still be used.

## Fishing mortality reference points

$\mathrm{F}_{0.1}$ and $\mathrm{F}_{\max }$ are estimated by the MFDP yield per recruit routine, and increased from 0.08 to 0.15 and from 0.14 to 0.30 for $\mathrm{F}_{0.1}$ and $\mathrm{F}_{\text {max }}$, respectively, in the $1999-2005$ assessments. In the 2010 assessment $F_{0.1}$ and $F_{\max }$ were estimated to 0.08 and 0.33 , respectively.

The values of Flow, Fmed and Fhigh obtained by the 2002 WG were $0.11,0.34$ and 0.69 , respectively.

The SGPAFM (ICES 1998/ACFM: 10) suggested the limit reference point Flim $=$ Fmed $^{\text {med }}$ for Northeast Arctic cod, haddock and saithe. A precautionary fishing mortality ( $\mathrm{F}_{\mathrm{pa}}$ ) was defined as $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\text {lim }} \cdot \mathrm{e}^{-1.645^{\sigma}}(\sigma=0.2-0.3)$. The 1998 WG , however, found that setting $\mathrm{F}_{\text {lim }}=\mathrm{F}_{\text {med }}$ did not correspond very well with the exploitation history for those fish stocks. It was therefore decided to estimate $\mathrm{F}_{\mathrm{pa}}$ and other reference points by the PASoft program package (MRAG 1997). The estimates for $\mathrm{F}_{0.1}, \mathrm{~F}_{\max }$, and $\mathrm{F}_{\text {med }}$ were exactly the same as the values already estimated by other routines. The median value for Floss was estimated at 0.43. Flim can be set at Floss (ICES 1998/ACFM:10). The probability of exceeding Flim should be no more than $5 \%$ (ICES 1997/Assess: 7). The $5^{\text {th }}$ percentile of the Floss estimated here was 0.30 and the 1998 WG recommended using this value for $F_{\text {pa. }}$ ACFM considered the $5^{\text {th }}$ percentile calculated from the PASoft program package to be too unstable for long term use and re-estimated $\mathrm{F}_{\mathrm{pa}}$ using the formula $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\text {lim }}$ $\cdot \mathrm{e}^{-1.645^{\sigma}}$ with $\sigma=0.3$ giving a $\mathrm{F}_{\mathrm{pa}}=0.26$, based on an estimated $\mathrm{F}_{\text {lim }}=0.45$ (ICES 1998c). An updated version of the PASoft program package (CEFAS 1999) was available at the 1999 WG and $\mathrm{F}_{\mathrm{pa}}$ was re-estimated to 0.26 . The WG therefore agreed to use this value for a precautionary fishing mortality for saithe ( $\mathrm{F}_{\mathrm{pa}}=0.26$ ).

ICES CM 2003/ACFM:15 proposed that $\mathbf{F}_{\text {lim }}$ should be set on the basis of $\mathbf{B}_{\mathrm{lim}}$, and $\mathbf{F}_{\text {lim }}$ should be derived deterministically as the fishing mortality that will on average (i.e. with a $50 \%$ probability) drive the stock to the biomass limit. The functional relationship between spawner-per-recruit and F will then give the F associated with the R/SSB slope derived from the $B_{\lim }$ estimate obtained from the segmented regression. At the 2005 WG arithmetic means of proportion mature 1960-2004, weight in stock and weight in catch 1980-2004 (weights were constant before 1980), natural mortality and fishing pattern 1960-2004 were used for calculating the spawner-per-recruit function using ICES Secretariat yield-per-recruit software. R/SSB = 1.27 from the $\mathbf{B}_{\text {lim }}$ esti-
mation gives $\mathrm{SSB} / \mathrm{R}=0.7874$ and a $\mathrm{F}_{\text {lim }}=0.58$. Applying the "magic formula" $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\lim }$ $\exp \left(-1.645^{*} \sigma\right)$, gives a $\mathrm{F}_{\mathrm{pa}}$ of 0.35 .

At the 2010 WG the latter procedure was repeated. Arithmetic means of proportion mature 1960-2009, weight in stock and weight in catch 1980-2009 (weights were constant before 1980), natural mortality and fishing pattern 1960-2009 w ere used for calculating the spawner-per-recruit function using ICES Secretariat yield-per-recruit software. $\mathrm{R} / \mathrm{SSB}=1.48$ from the $\mathrm{B}_{\mathrm{lim}}$ estimation gives $\mathrm{SSB} / \mathrm{R}=0.676$ and a $\mathrm{F}_{\mathrm{lim}}=0.59$. Applying the "magic formula" $\mathrm{F}_{\mathrm{pa}}=\mathrm{F}_{\mathrm{lim}} \exp \left(-1.645^{*} \sigma\right)$, gives a $\mathrm{F}_{\mathrm{pa}}$ of 0.36 . As explained above, the existing values of $\mathrm{Flim}_{\mathrm{l}}=0.58$ and $\mathrm{F}_{\mathrm{pa}}=0.35$ will still be used.

## H. Other Issues

## Harvest control rule

In 2007 Norway asked ICES to evaluate whether a proposal for a harvest control rule for setting the annual fishing quota (TAC) for Northeast Arctic saithe was consistent with the precautionary approach. The harvest control rule contains the following elements:

- estimate the average TAC level for the coming 3 years based on $\mathbf{F}_{\text {pa. }}$ TAC for the next year will be set to this level as a starting value for the 3-year period.
- the year after, the TAC calculation for the next 3 years is repeated based on the updated information about the stock development. However, the TAC should not be changed by more than $+/-15 \%$ compared with the previous year's TAC.
- if the spawning stock biomass (SSB) in the beginning of the year for which the quota is set (first year of prediction), is below $\mathbf{B}_{\mathrm{pa}}$, the procedure for establishing TAC should be based on a fishing mortality that is linearly reduced from $\mathbf{F}_{\mathrm{pa}}$ at $\mathrm{SSB}=\mathbf{B}_{\mathrm{pa}}$ to 0 at SSB equal to zero. At SSB levels below $\mathbf{B}_{\mathrm{pa}}$ in any of the operational years (current year and 3 years of prediction) there should be no limitations on the year-to-year variations in TAC.

ICES concluded that the HCR is consistent with the precautionary approach for all simulated data and settings, including a rebuilding situation under the condition that the assessment uncertainty and error are not greater than those calculated from historic data (ICES 2007). This also holds true when an implementation error (difference between TAC and catch) equal to the historic level of $3 \%$ is included.

The highest long-term yield was obtained for an exploitation level of 0.32, i.e. a little below the target F used in the HCR (Fpa), and ICES recommended using a lower value in the HCR.

The HCR is expected to rebuild a depleted stock to a level above Blim within three years.

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Stock: Sebastes mentella (Beaked Redfish) in Subareas I and II<br>Working Group: Arctic Fisheries Working Group (AFWG)<br>Date: 06.05.10

## A. General

## A.1. Stock definition

The stock of Sebastes mentella (beaked redfish) in ICES Subareas I and II, also called the Norwegian-Barents Sea stock, is found in the northeast Arctic from $62^{\circ} \mathrm{N}$ in the south to the Arctic ice north and east of Spitsbergen. The south-western Barents Sea and the Spitsbergen areas are first of all nursery areas. Although some adult fish may be found in smaller subareas, the main behaviour of S. mentella is to migrate westwards and south-westwards towards the continental slope and out in the pelagic Norw egian Sea as it grows and becomes adult. In the Norw egian Sea and along the slope south of $70^{\circ} \mathrm{N}$ only few specimens less than 28 cm are observed, and on the shelf south of this latitude S. mentella are only found along the slope from about 450 m down to about 650 m depth. The southern limit of its distribution is not well defined but is believed to be somewhere on the slope northwest of Shetland. The stock boundary $62^{\circ} \mathrm{N}$ is therefore more for management purposes than a biological basis for stock separation, although the abundance of this species south of this latitude becomes less. The main areas of larval extrusion are along the slope from north of Shetland to west of Bear Island. The peak of larval extrusion takes place during the first half of April. Genetic studies have not revealed any hybridisation with S. marinus or S. viviparus in the area. Recent genetic studies revealed no differentiation between $S$. mentella in the Norwegian Sea and the Barents Sea.

## A.2. Fishery

The only directed fisheries for Sebastes mentella (deep-sea redfish) are trawl fisheries. By-catches are taken in the cod fishery and as juveniles in the shrimp trawl fisheries. Traditionally, the fishery for S. mentella was conducted by Russia and other East European countries on grounds located south of Bear Island towards Spitsbergen. The highest landings of S. mentella were $269,000 \mathrm{t}$ in 1976. This was followed by a rapid decline to $80,000 \mathrm{t}$ in 1980-1981 then a second peak of $115,000 \mathrm{t}$ in 1982. The fishery in the Barents Sea decreased in the mid-1980s to the low level of $10,500 \mathrm{t}$ in 1987. At this time Norwegian trawlers showed interest in fishing S. mentella and started fishing further south, along the continental slope at approximately 500 m depth. These grounds had never been harvested before and were inhabited primarily by mature redfish. After an increase to $49,000 \mathrm{t}$ in 1991 due to this new fishery,
landings have been at a level of $10,000-15,000 \mathrm{t}$, except in 1996-1997 when they dropped to $8,000 \mathrm{t}$. Since 1991 the fishery has been dominated by Norway and Russia. Since 1997 ACFM has advised that there should be no directed fishery and that the by-catch should be reduced to the lowest possible level.

The redfish population in Subarea IV (North Sea) is believed to belong to the North-east Arctic stock. Since this area is outside the traditional areas handled by this Working Group, the catches are not included in the assessment. The landings from Subarea IV have been 1,000-3,000 t per year. Historically, these landings have been S. marinus, but since the mid-1980s trawlers have also caught S. mentella in Subarea IV along the northern slope of the North Sea. Approximately $80 \%$ of the Norwegian catches are considered to be S. mentella.

Strong regulations were enforced in the fishery in 1997. Since then it has been forbidden to fish redfish (both S.marinus and S. mentella) in the Norwegian EEZ north and west of straight lines through the positions:

1. N 7000' E 0521'
2. N 7000' E 1730'
3. N 7330' E 1800'
4. N 7330' E 3556'
and in the Svalbard area (Division IIb). When fishing for other species in these areas, a maximum $25 \%$ by-catch (in weight) of redfish in each trawlhaul is allowed.

To provide additional protection of the adult S. mentella stock, two areas south of Lofoten have been closed for all trawl fishing since 1 March 2000. The two areas (A and B) are delineated by straight lines between the following positions:

| A | B |
| :---: | :---: |
| 1. N 6630' E 0659' | 1.N 6236' E 0300' |
| 2. N 6621' E 0644' | 2.N 6210' E 0115' |
| 3.N 6543' E 0600' | 3.N 6240' E 0052' |
| 4.N 6520' E 0600' | 4.N 6300' E 0300' |
| 5.N 6520' E 0530' |  |
| 6. N 6600' E 0530' |  |

7. N 6630' E 0634.27'

Area A has recently been enlarged to include the continental slope north to $\mathrm{N} 67^{\circ} 10^{\prime}$.
Since 1 January 2003 all directed trawl fishery for redfish (both S. marinus and $S$. mentella) is forbidden in the Norwegian Economic Zone north of $62^{\circ} \mathrm{N}$. When fishing for other species it is legal to have up to $20 \%$ redfish (both species together) in round weight as bycatch per haul and on board at any time. Since 1 January 2005 the bycatch percentage has been reduced to $15 \%$ (both species together).

From 1 January 2000 until 31 December 2005 a maximum legal by-catch criterion of 10 juvenile redfish (both S.marinus, S. mentella and S. viviparus) per 10 kg shrimp has been enforced in the shrimp fishery. Since 1 January 2006 this by-catch criterion has been reduced to 3 juvenile redfish (both S.marinus, S. mentella and S. viviparus) per 10 kg shrimp.

Landings of $S$. mentella taken in the pelagic fishery for blue whiting and herring in the Norw egian Sea have for some countries for some years been reported to the working group. In 2004-2006 this fishery developed further to become a directed and free fishery in 2006. Since 2007 NEAFC has decided on a TAC to be fished in an olympic fishery. In 2008, seven countries and 31 trawlers were involved in this fishery. Although sporadic registrations and scattered catches of $S$. marinus may be observed, biological samples of the catches collected by observers and fishers show that the commercial catches are completely dominated by the deep-water redfish S. mentella.
Vinnichenko (WD9, AFWG 2007) gives a good and comprehensive description of the previous abundance of pelagic $S$. mentella in the international waters of the Norw egian Sea, and how by-catches and exploratory fishing have developed during 1979-2006. According to Vinnichenko, in 1998-2000 small by-catches of redfish (no more than 8 t per year) were reported from the blue whiting and herring fisheries in the international waters of the Norwegian Sea and in the Norw egian Economic Zone. In 2001-2003 occurrence of redfish was reported from a larger area and catches increased to 60-118 t.

In 2004 the amount of redfish in catches increased significantly, and in June-August this species was more frequently occurring in the south of the sea. In September catches of redfish ( 0.5 t per hour haul) were reported from international waters and the NEZ. In October, in the northern part of the international waters, trawlers had a catch of redfish of $0.5-10 \mathrm{t}$ per day, sometimes to $15-40 \mathrm{t}$. By-catches of redfish were also reported from the Bear Island-Spitsbergen area and the NEZ. The total reported catch of pelagic S. mentella in 2004 was 1,512 t.

In summer of 2005 small quantities of redfish were steadily present in catches on the blue whiting and herring fisheries in the international waters of the Norwegian Sea and the Bear Island-Spitsbergen area. In the first half of September some vessels operating in the Bear Island-Spitsbergen reported by-catches of $S$. mentella as large as 6-25 t per day. In the end of September in the north of the international waters of the Norwegian Sea large Russian trawlers for the first time began fishing for redfish in a directed fishery. They fished with a gigantic "Gloria" trawl. The fishery finished in the beginning of November after the redfish dispersed. In 2005 the Russian fleet reported a catch of S. mentella of 3299 t , including the by-catch in the blue whiting and herring fisheries. Fishing for redfish was also conducted by a Faroese trawler. Besides, small quantities of redfish were fished by German vessels in the blue whiting fishery.
In 2006 first small catches of redfish (to 50 kg per haul) were reported from the herring fishery in the NEZ in February. In June-August catches of redfish of 70-120 kg per hour haul were reported in the blue whiting and mackerel fisheries in the international waters south of $70^{\circ} \mathrm{N}$. Targeted redfish fishery by the Faroese and Russian trawlers began at the Mona Ridge (i.e., the ridge separating the Norw egian Sea into two main basins) in August. By mid-September the number of fishing vessels operating in that area was as high as 40 vessels, including 8-12 vessels from Russia and up to 30 vessels from Iceland, Faroe Islands, Norway and EU. In October 15-25 vessels continued the fishery. It finished in mid-November as the fish then had disappeared from the area. The Russian catch in the directed S. mentella fishery was 9,157 t. Redfish also occurred in catches by trawlers, that fished for blue whiting and herring. The total reported catch of pelagic S. mentella by Russian vessels in 2006 was $9,390 \mathrm{t}$, and a total of $28,429 \mathrm{t}$ by all nations during this non-regulated fishery in 2006.

For 2007, the North East Atlantic Fisheries Commission (NEAFC) agreed to set a TAC of 15500 t that could be fished in international waters in an olympic fishery (i.e., free competition among vessels until the TAC is taken) starting on 1 September. Information about the fishery in 2007 was presented to the working group in 2008 by several countries. A total catch of 15808 t S. mentella has been reported to ICES and the AFWG, as caught in the pelagic fisheries in the Norwegian Sea, incl. minor bycatches in the blue whiting and herring fisheries.
For 2008, the North East Atlantic Fisheries Commission (NEAFC) agreed to set a TAC of $14,500 \mathrm{t}$ that could be fished in international waters in an olympic fishery starting on 1 September. Only Portugal provided a Working Document about this fishery (WD 2), but in addition, Russia and Spain, provided length distribution of their pelagic catches. Norway distributed their pelagic catches by length and age using data collected during the scientific survey in the fishing area one week before the fishing started. A total catch of $9,183 \mathrm{t}$ S. mentella has been reported to ICES and the AFWG as caught in the pelagic fisheries in the Norw egian Sea.
In 2009, NEAFC set a TAC of 10,500 t that could be fished in international waters in an olympic fishery starting on $15^{\text {th }}$ August. Preliminary figures indicate that a total catch of only $5,291 \mathrm{t}$ was reported to NEAFC for the pelagic fishery in that year.
Some countries have only reported catches taken in Sub-area IIa, without information whether the fish were caught pelagic or demersal. For these countries, the WG has considered all catches not reported to Norwegian authorities as being caught in international waters outside the EEZ.

Bycatch of herring could be a problem during day-time trawling in these waters at this time of the year. In some catches with the research survey trawl $(40 \mathrm{~mm}$ mesh size in codend) up to $30 \%$ (in weight) herring was caught as bycatch when targetting the redfish. Even with a commercial trawl ( 100 mm mesh size in codend) reports from the fishery show that mixed catches of herring may happen. Even if some of the herring is selected out through the meshes, mortality through mesh selection may be high. During the 2007 olympic fishery bycatches of blue whiting were small. Best catch-rates of S. mentella were usually done during day-time. According to the skippers they observed and got the best catch-rates of redfish about 50 meters deeper than last year, i.e. at about 400 m . Two tons redfish per trawl hour was considered as a very good catch rate. With a common haul duration of 18 hours, catch rates of $30-40$ tons/day were not uncommon. Even catch rates up to 70 tons/day were reported.

## A.3. Ecosystem aspect

As 0-group and juvenile this stock is an important plankton eater in the Barents Sea, and when this stock was sound, 0 -group were observed in great abundance in the upper layers utilizing the plankton production. Especially during the first five-six years of life S. mentella is also preyed upon by other species, of which its contribution to the cod diet is well documented.

## B. Data

## B.1. Commercial catch

The landings statistics used by the Arctic Fisheries W orking Group (AFWG) are those officially reported to ICES. In cases where such reportings to ICES do not exist, reportings made directly to Norwegian authorities during the fishery have been used
as preliminary figures. Norwegian commercial catch in tonnes by quarter, area and gear are derived from the sales notes statistics of The Directorate of Fisheries. Data are aggregated on 17 areas for bottom trawl. For bottom trawl the quarterly area distribution of the catches is area adjusted by logbook data from The Directorate of Fisheries. No discards are reported or accounted for. Reliable estimates of species breakdown (S. mentella vs. S. marinus) by area are available back to 1989. The national landings of redfish for Norway and Russia are split into species by the respective national laboratories. For other countries (and areas) the AFWG has split the landings into S. mentella and S. marinus based on reports from different fleets to the Norw egian fisheries authorities.

The Norwegian sampling strategy is to have age-length samples from all major gears in each area and quarter. There are at present no defined criteria on how to allocate samples of catch numbers, mean length and mean weight at age to unsampled catches, but the following general process has been applied: First look for samples from a neighbouring area if the fishery extends to this area in the same quarter. If there are no samples available in neighbouring areas, search in neighbouring quarters, first from the same gear in the same area, and than from neighbouring areas and similar gears. The last option is to search for samples from other gears with the most similar selectivity in the same area or in neighbouring areas. For some gears, areas and quarters length samples taken by the coast guard are applied and combined with an ALK from a neighbouring area, gear or quarter. ALKs from research surveys (shrimp trawl) are also used to fill holes.

For Norway, weights at age in the catch are estimated according to the formula which gives the best fit to the length-weight data pairs collected during the year and applied to the mean length at age

The text table below shows which country supplies which kind of data:

|  | Kind of data |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Country | Caton (catch in weight) on unidentified redfish | Caton (catch in weight) on S. mentella | Canum <br> (catch at age in numbers) | Weca (weight at age in the catch) | Matprop (proportion mature by age) | Length composition in catch |
| Norway |  | X | x | x |  | X |
| Russia |  | X | $\mathrm{x}^{2)}$ | $\mathrm{x}^{2)}$ | x (86-01) | X |
| Germany | X | $\mathrm{x}^{3)}$ |  |  |  | $\mathrm{x}^{3)}$ |
| United Kingdom | X | 1) |  |  |  |  |
| France | X | 1) |  |  |  |  |
| Spain | X | 1) |  |  |  |  |
| Portugal | X | 1) |  |  |  |  |
| Ireland | X | 1) |  |  |  |  |
| Greenland | X | 1) |  |  |  |  |
| Faroe Islands ${ }^{1}$ |  |  |  |  |  |  |
| Iceland | X | 1) |  |  |  |  |

${ }^{1)}$ As reported to Norwegian authorities during the fishery (only for the Norwegian Economic Zone and Svalbard)
${ }^{2)}$ For main fishing area until 2001
${ }^{3)}$ Irregularly

The Norwegian, Russian and German input files are Excel spreadsheet files. The data should be found in the national laboratories and with the stock co-ordinator. The data will soon be included in InterCatch

The national data have been aggregated to international data on Excel spreadsheet files. The Russian and German length composition has been applied on the Russian
and German landings, respectively, using an age-length-key (ALK) and weight at age data from the Norwegian trawl landings. Catches from the other countries were assumed to have the same age composition and weight at age as the Norwegian trawl landings. In some years the final German and Russian numbers at age have been adjusted to remove SOP discrepancies before aggregation to international data. The Excel spreadsheet files used for age distribution, adjustments and aggregations can be found with the stock co-ordinator and for the current and previous year in the ICES AFWG Sharepoint under 'Data'.
Historic result files (FAD data) can be found at ICES and with the stock co-ordinator, either in the IFAP system as SAS datasets or as ASCII files on the Lowestoft format, either under w:lacfmlafwgl<year>datalsmn_arct or w:lifapdataleximportlafwglsmn_arct.

## B.2. Biological

Since 1991, the catch in numbers at age of S. mentella from Russia is based on otolith readings. The Norw egian catch-at-age is based on otoliths back to 1990. Before 1990, when the Norwegian catches of S. mentella were smaller, Russian scalebased agelength keys were used to convert the Norw egian length distribution to age.

As input to trial analytical assessments, weight at age in the stock is assumed to be the same as weight at age in the catch.

A fixed natural mortality of 0.1 is used both in the assessment and the forecast.
Both the proportion of natural mortality before spawning (Mprop) and the proportion of fishing mortality before spawning (Fprop) are set to 0 .

Age-based maturity ogives for S. mentella (sexes combined) are available for 19861993, 1995 and 1997-2001 from Russian research vessel observations in spring. Average ogives for 1966-1972 and 1975-1983 havebeen used for the periods 1965-1975 and 1976-1983, respectively. Average ogives for 1975-1983, 1984-1985 and data for 1986-1993 (Table D8) were used to generate a smoothed maturity ogive for 1984-1992 (3 year running average). The 1992-1993 average was used for 1993 and 1994, the 1995 data for 1995, the average for 1995 and 1997 for 1996, and the collected material for the subsequent years up to 2001 were taken as representative for these years.

## B.3. Surveys

The results from the following research vessel survey series have annually been evaluated by the AFWG:

1) The international 0-group survey (since 2004 part of the Ecosystem survey) in the Svalbard and Barents Sea areas in August-September since 1980 (incl.).
2) Russian bottom trawl survey in the Svalbard and Barents Sea areas in October-December since 1978 (incl.) in fishing depths of 100-900 m.
3) Norwegian Svalbard (Division IIb) bottom trawl survey (AugustSeptember) since 1986 (incl.) in fishing depths of $100-500 \mathrm{~m}$. Data disaggregated on age only since1992.
4 ) Norw egian Barents Sea bottom trawl survey (February) since 1986 (incl.) in fishing depths of $100-500 \mathrm{~m}$. Data disaggregated on age only since 1992.

Although the Norwegian Svalbard (August-September) and Barents Sea (February) groundfish surveys are conducted at different times of the year and may overlap in the south of Bear Island area, the two series can be combined to get an approximate total estimate for the whole area.

1) The Norwegian survey initially designed for redfish and Greenland halibut is now part of the ecosystem survey and covers the Norwegian Economic Zone (NEZ) and Svalbard incl. north and east of Spitsbergen during August 1996-2008 from less than 100 m to 800 m depth. This survey includes survey no. 3 above, and has been a joint survey with Russia since 2003, and since then called the Ecosystem survey.
2 ) Russian acoustic survey in April-May since 1992 (except 1994, 1996 and 2002-2004) on spawning grounds in the western Barents Sea .

The international 0-group fish survey carried out in the Barents Sea in AugustSeptember since 1965 does not distinguish between the species of redfish but it is believed to be mostly S. mentella. The survey design has improved and the indices earlier than 1980 are not directly comparable with subsequent years.

Russian acoustic surveys estimating the commercially sized and mature part of the $S$. mentella stock have been conducted in April-May on the Malangen, Kopytov, and Bear Island Banks since 1986. In 1992 the area covered was extended, and data on age are available for 1992-1993, 1995 and 1997-2001. This is the only survey targeting commercially sized S. mentella, but only a limited area of its distribution.

In order to investigate the distribution and abundance of pelagic Sebastes mentella in the Norwegian Sea the following surveys are/have been conducted:
i. Norw egian part of the international ecosystem survey in the Nordic Seas in spring 2007-2009 (PGNAPES).
ii. Norwegian trawl and acoustic survey in September 2007, and ICES coordinated international trawl and acoustic survey conducted by Norway, Russia and the Faroes in August 2008.

## B.4. Commercial CPUE

Revised catch-per-hour-trawling data for the S. mentella fishery have been available from Russian PST- and BMRT-trawlers fishing in ICES Division IIa in March-May 1975-2002, representative for the directed Russian fishery accounting for $60-80 \%$ of the total Russian catch. The W orking Group mean that the Russian trawl CPUE series do not represent the trend in stock size but is more a reflection of stock density. This is because the fishery on which these data are based since 1996 was carried out by one or two vessels on localised concentrations in the Kopytov area southwest of Bear Island. This is also reflected by the relative low effort at present. Due to this change in fishing behaviour/effort, CPUEs have been plotted only for the period after 1991.

## B.5. Other relevant data

None

## C. Historical Stock Development

Model used:
Software used:
Model Options chosen:
Input data ty pes and characteristics:

| TYPE | NAME | Year Rang E | AGERANGE | VARIABLEFROM YEAR TO YEAR YES/NO |
| :---: | :---: | :---: | :---: | :---: |
| Caton | Catch in tonnes | 1965-2008 | 6-19+ | yes |
| Canum | Catch at age in numbers | 1965-2008 ${ }^{1}$ | 6-19+ | yes |
| Weca | Weight at age in the commercial catch | 1965-2008 | 6-19+ | yes |
| West | Weight at age of the spawning stock at spawning time. | 1965-2008 | 6-19+ | yes |
| Mprop | Proportion of natural mortality before spawning | 1965-2008 | 6-19+ | Constant $=0$ |
| Fprop | Proportion of fishing mortality before spawning | 1965-2008 | 6-19+ | Constant $=0$ |
| Matprop | Proportion mature at age | 1965-2008 | 6-19+ | $\begin{aligned} & 1965-1975 \text {, const. } \\ & 1976-1983 \text {, const. } \\ & \text { 1984-2001,variable } \\ & 2002-\text {, const } \end{aligned}$ |
| Natmor | Natural mortality | 1965-2008 | 6-19+ | Constant $=0.1$ |

${ }^{1}$ Based on otoliths since 1991
Tuning data: files not updated since 2005, but data/results exist also for recent years

| TYPE | NAME | YEAR RANGE | AGERANGE |
| :--- | :--- | :--- | :--- |
| Tuning fleet 1 | FLT10 Rus young | $1991-2005$ | $6-8$ |
| Tuning fleet 2 | FLT13 Rus acous | $1995-2001$ | $6-14$ |
| Tuning fleet 3 | FLT14 Norw bottom | $1996-2005$ | $2-11$ |
| $\ldots$. |  |  |  |

## D. Short-Term Projection

Model used: Visual analysis of survey results.
Software used: none
Initial stock size:
Maturity:
F and Mbefore spawning:
Weight at age in the stock:
Weight at age in the catch:
Exploitation pattern:
Intermediate year assumptions:
Stock recruitment model used:

Procedures used for splitting projected catches:

## E. Medium-Term Projections

Model used: Visual analysis of survey results.
Software used: none
Initial stock size:
Natural mortality:
Maturity:
F and Mbefore spawning:
Weight at age in the stock:
Weight at age in the catch:
Exploitation pattern:
Intermediate year assumptions:
Stock recruitment model used:
Uncertainty models used:

1. Initial stock size:
2. Natural mortality:
3. Maturity:
4. F and Mbefore spawning:
5. Weight at age in the stock:
6. Weight at age in the catch:
7. Exploitation pattern:
8. Intermediate year assumptions:
9. Stock recruitment model used:

## F. Long-Term Projections

Model used:
Software used:
Maturity:
F and Mbefore spawning:
Weight at age in the stock:
Weight at age in the catch:
Exploitation pattern:
Procedures used for splitting projected catches:
G. Biological Reference Points
H. Other Issues
I. References

Stock specific documentation of standard assessment procedures used by ICES.

| Stock:... | Golden redfish Sebastes marinus in ICES <br> Subareas I and II |
| :--- | :--- |
| Working Group | Arctic Fisheries Working Group |
| Date: | 06.05 .2010 |

## A. General

## A.1. Stock definition

The stock of Sebastes marinus (golden redfish) in ICES Subareas I and II is found in the northeast Arctic from $62^{\circ} \mathrm{N}$ in the south to north of Spitsbergen. The Barents Sea area is first of all a nursery areas, and relatively few fish are distributed outside Spitsbergen. S. marinus are distributed all over the continental shelf southwards to beyond $62^{\circ} \mathrm{N}$, and also along the coast and in the fjords. The main areas of larval extrusion are outside Vesteralen, on the Halten Bank area and on the banks outside Mere. The peak of larval extrusion takes place ca. one month later than S. mentella, i.e. during beginning of May. Genetic studies have not revealed any hybridisation with S. marinus or S. viviparus in the area.

## A.2. Fishery

The fishery for Sebastes marinus (golden redfish) is mainly conducted by Norway which accounts for $80-90 \%$ of the total catch. Germany also has a long tradition of a trawl fishery for this species. The fish are caught mainly by trawl and gillnet, and to a lesser extent by longline and handline. The trawl and gillnet fishery have benefited from the females concentrating on the "spawning" grounds during spring. Some of the catches, and most of the catches taken by other countries, are taken in mixed fisheries together with saithe and cod. Important fishing grounds are the Møre area (Svinøy), Halten Bank, the banks outside Lofoten and Vesterålen, and Sleppen outside Finnmark. Traditionally, S. marinus has been the most popular and highest priced redfish species.

Until 1 January 2003 there were no regulations particular for the S. marinus fishery, and the regulations aimed at $S$. mentella had only marginal effects on the $S$. marinus stock. After this date, all directed trawl fishery for redfish (both S. marinus and S. mentella) is forbidden in the Norwegian Economic Zone north of $62^{\circ} \mathrm{N}$. During 2003 and 2004, when fishing for other species it was legal to have up to $20 \%$ redfish (both species together) in round weight as bycatch per haul and on board at any time. Since 1 January 2005 this percentage has been reduced to $15 \%$.
A minimum legal catch size of 32 cm has been set for all fisheries (since 14 April 2004), with the allowance to have up to $10 \%$ undersized (i.e., less than 32 cm ) specimens of S.marinus (in numbers) per haul.

Until April 2004 there were no regulations of the other gears/fleets than trawl fishing for $S$. marinus. Since then, different limited moratoriums have been enforced in all fisheries except trawl and handline vessels less than 11 meters. The moratorium has been from 1-31 May in 2004, 20 April-19 June in 2005 and during April-May and September in 2006. Since 2007 the moratorium has been during 5 months, i.e., MarchJune and September. When fishing for other species (also during the moratorium) it is allowed for these fleets to have up to $15 \%$ (in 2004, 20\%) by catch of redfish (in round weight) summarized during a week fishery from Monday to Sunday.
Since 1 January 2006 it is forbidden to use gillnets with meshsize less than 120 mm when fishing for redfish.

Since 1 January 2006, the maximum bycatch of redfish (both S. mentella and S. mari$n u s$ ) juveniles in the international shrimp fisheries in the northeast Arctic has been reduced from ten to three redfish per 10 kg shrimp.

## A.3. Ecosystem aspects

## B. Data

## B.1. Commercial catch

The landings statistics used by the Arctic Fisheries W orking Group (AFWG) are those officially reported to ICES. In cases where such reportings to ICES do not exist, reportings made directly to Norwegian authorities during the fishery have been used as preliminary figures. Norwegian commercial catch in tonnes by quarter, area and gear are derived from the sales notes statistics of The Directorate of Fisheries. Data from about 20 sub areas are aggregated for the gears gill net, long line, hand line, Danish seine and bottom trawl. For bottom trawl the quarterly area distribution of the catches is area adjusted by logbook data from The Directorate of Fisheries. No discards are reported or accounted for. Reliable estimates of species breakdown (S. mentella vs. S. marinus) by area are available back to 1989. The national landings of redfish for Norway and Russia are split into species by the respective national laboratories. For other countries (and areas) the AFWG has split the landings into $S$. mentella and S. marinus based on reports from different fleets to the Norwegian fisheries authorities.

The Norwegian sampling strategy is to have age-length samples from all major gears in each area and quarter. There are at present no defined criteria on how to allocate samples of catch numbers, mean length and mean weight at age to unsampled catches, but the following general process has been applied: First look for samples from a neighbouring area if the fishery extends to this area in the same quarter. If there are no samples available in neighbouring areas, search in neighbouring quarters, first from the same gear in the same area, and then from neighbouring areas and similar gears. The last option is to search for samples from other gears with the most similar selectivity in the same area or in neighbouring areas. For some gears, areas and quarters length samples taken by the coast guard are applied and combined with an ALK from a neighbouring area, gear or quarter. ALKs from research surveys (shrimp trawl) are also used to fill holes.

For Norway, weights at age in the catch are estimated according to the formula which gives the best fit to the length-w eight data pairs collected during the year and applied to the mean length at age.

The text table below shows which country supplies which kind of data:

|  | Kind of data |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Country | Caton (catch in weight) on unidentified redfish | Caton (catch in weight) on S. marinus | Canum (catch at age in numbers) | Weca (weight at age in the catch) | Matprop (proportion mature by age) | Leng th composition in catch |
| Norway |  | x | x | x |  | x |
| Russia |  | x |  |  |  | x |
| Germany | x | $\mathrm{x}^{2}$ |  |  |  | x |
| United | x | ${ }^{1)}$ |  |  |  |  |
| King dom | x | ${ }^{1)}$ |  |  |  |  |
| France | x | 1) |  |  |  |  |
| Spain | x | 1) |  |  |  |  |
| Portugal | x | 1) |  |  |  |  |
| Ireland | x | 1) |  |  |  |  |
| Greenland |  |  |  |  |  |  |
| Faroe Islands ${ }^{1)}$ | x | 1) |  |  |  |  |
| Iceland |  |  |  |  |  |  |

${ }^{1)}$ As reported to Norwegian authorities during the fishery (only for the Norwegian Economic Zone and Svalbard)
${ }^{2)}$ Irregularly

The Norwegian and German input files are Excel spreadsheet files, while the Russian input data are supplied on paper and later punched into Excel spreadsheet files before aggregation to international data. The data should be found in the national laboratories and with the stock co-ordinator.

The national data have been aggregated to international data on Excel spreadsheet files. The Russian and German length composition has been applied on the Russian and German landings, respectively, using an age-length-key (ALK) and weight at age data from the Norwegian trawl landings. Catches from the other countries were assumed to have the same age composition and weight at age as the Norwegian trawl landings. In some years the final German and Russian numbers at age have been adjusted to remove SOP discrepancies before aggregation to international data. The Excel spreadsheet files used for age distribution, adjustments and aggregations can be found with the Norwegian stock co-ordinator and for the current and previous year in the ICES computer system under w:lacfmlafwgl<year>lpersonalname (of stock co-ordinator).
The result files (FAD data) can be found at ICES and with the stock co-ordinator, either in the IFAP system as SAS datasets or as ASCII files on the Lowestoft format, either under $\mathbf{w}$ :lacfmlafwg\<year>\datalsmr-arct or w: $\mathbf{~ l i f a p d a t a l e x i m p o r t l a f w g | s m r - ~}$ arct.

## B.2. Biological

The total catch-at-age data back to 1991 are based on Norwegian otolith readings. In 1989-1990 it was a combination of the German scale readings on the German catches, and Norwegian otolith readings for the rest. In 1984-1989 only German scale readings were available, while in the years prior to 1984 Russian scale readings exist.

Weight at age in the stock is assumed to be the same as weight at age in the catch.
When an analytical assessment is made, a fixed natural mortality of 0.1 is used both in the assessment and the forecast.

Both the proportion of natural mortality before spawning (Mprop) and the proportion of fishing mortality before spawning (Fprop) are set to 0 .

A knife-edge maturity at age 15 (age 15 as $100 \%$ mature) has been used for this stock. Since 2006 a maturity ogive has been modelled and estimated by the GADGET model.

## B.3. Surveys

The results from the following research vessel survey series have annually been evaluated by the Working Group:

1 ) Norwegian Barents Sea bottom trawl survey (February) from 1986-2009 in fishing depths of 100-500 m. Data are available on length for the years 1986-2009, and on age for the years 1992-2008. This survey covers important nursery areas for the stock
2 ) Norwegian Svalbard (Division IIb) bottom trawl survey (AugustSeptember) from 1985-2008 in fishing depths of 100-500 m. This survey covers the northernmost part of the species' distribution.

3 ) Data on length and age from both these surveys have been simply added together and used in the assessments.
4 ) Catch rates (numbers/nautical mile) and acoustic indices of Sebastes marinus from the Norwegian Coastal and Fjord survey in 1995-2008 from Finnmark to Møre. Since 2003, only catch rates are available.

## B.4. Commercial CPUE

The former (until 2002) CPUE-series for $S$. marinus from Norwegian 32-50 meter freezer trawlers has been improved (e.g., analysing the trawl data with regards to vessel length instead of vessel tonnage) and presented from 1992 onwards. Only data from days with more than $10 \%$ S. marinus in the catches (in weight) were included in the annual averages together with data on vessel days (i.e., effort) meeting the $10 \%$ criterion.

## B.5. Other relevant data

None.

## C. Historical Stock Development

The development of the stock has annually been discussed and evaluated based on the research survey series, and information from the fishery.

In some years trial analytical XSA assessments have been made and discussed by the Working Group.
Since WG2005, experimental analytical assessments have been conducted on this stock using GADGET, and results presented for the years 1990 - last year.

The GADGET model used for the assessment of S. marinus in areas I and II is closely related to the GADGET model that currently is used by the ICES North-Western WG on S. marinus (Björnsson and Sigurdsson 2003). The functioning of a Gadget model,
including parameter estimation, is described in Bogstad et al. (2004). The model used on this stock was for the first time presented to ACFM in 2005. The method was more thoroughly reviewed and described in AFWG report 2006. The main model period has been considered to be from 1990, with earlier years acting as a lead-in period to the model. S. marinus has been modelled with a single-species, single-area model, with mature and immature fish considered as two population groups. The fish were modelled in 1 cm length categories. The age and length ranges were defined as $3-30+$ and $1-59+\mathrm{cm}$, respectively.
S. marinus was considered to have Von Bertanlanffy growth (Nedreaas 1990) with parameters estimated within the model. The length-weight relationship $\mathrm{w}=0.000015^{*} \mathrm{l}^{\wedge} 3.0$ (where w is in kilogram and l in cm ) was used and kept constant between seasons and years. There has been no cannibalism or modelled predation mortality has been exclusively due to fishing and residual natural mortality was set initially at 0.1 . Recruitment was handled as a number of recruits estimated per year, and no attempt at closure of the life cycle was attempted. Maturity is explicitly modelled, allowing for a direct estimate of the spawning stock. Estimated parameters were: an L50 and slope parameters for the fleets, two growth parameters, annual recruitment, four parameters governing commercial selectivity (two per fleet), several parameters per survey governing selectivity (two per fleet), initial population numbers for mature and immature fish by age.

Data used for tuning are:

- Quarterly length distribution of the landings from two commercial fishing fleets
- Quarterly age-length keys from the same fishing fleets
- Length disaggregated survey indices from the Barents Sea (Division IIa) bottom trawl survey (February) from 1990-2009 (Table D12a).
- Age-length keys from the same survey (Table D12b).
- Length disaggregated catch rates (numbers/nautical mile) of Sebastes marinus from the Norwegian Coastal and Fjord survey in 1995-2008 from Finnmark to Møre (Division IIa)

The fishing was handled as two main, and two subsidiary fleets. The Norwegian trawl- and gillnet fleets were both fully modelled, with estimated selectivity for each, accounting for about $70-80 \%$ of the total catch in tonnes. The amount fished in each time step of one quarter of the year was input from catch data as a fixed amount. No account of possible errors in the catch-in-tons data was made. Two additional fleets have been considered; the international trawl fleet and a fleet made up by combining all other minor Norwegian fishing methods. Both these fleets have quarterly catch-intons specified, and have used the same selectivity as the Norwegian trawl fleet. In addition to catch-in-tons, quarterly catch-in-numbers-at-length and age-length keys have been used. The format of the selectivity (L50) was selected and assumed to remain constant over time for each fleet.

The Barents Sea survey data were used as age-length keys giving the distribution within a single year, and as a purely length based survey index giving year to year variations in numbers by length. Prior to 1992 only length and weight data were recorded; after that data on annual age readings (and hence age-length data) are also available. The time period 1990-2006 was used, and the age-length key for 1992 was also used as age-length key for 1990-1991.

## D. Short-Term Projection

Model used: Visual inspection/analysis of survey results together with information from the fishery and Gadget model outputs. No analytical short-term projection has been made for this stock.

## E. Medium-Term Projections

Model used: Visual inspection/analysis of survey results together with information from the fishery and Gadget model outputs. No analytical short-term projection has been made for this stock.

Uncertainty models used: None

## F. Long-Term Projections

Not done

## G. Biological Reference Points

Until an analytical assessment can be accepted and used as basis for reference points calculations for this stock, candidate reference points for the biomass could be set at the average biomass level, or at a certain percentage of this level, estimated by the Russian and Norwegian trawl surveys since 1986. ACFM is supporting this suggestions and states that U-type reference points could be developed provided that a sufficient long time series demonstrating a dynamic range is available. Also the reference point should be expressed in biomass units (SSB or fishable stock), and work has hence been initiated to present the survey time series also in biomass units (also as SSB and fishable stock).

A maximum exploitation rate of $5 \%$ has been suggested sustainable for long lived species like Sebastes spp. when the stocks show no sign of reduced reproductive potential (ref. pelagic redfish in the Irminger Sea and for several rockfishes in the Pacific). Based on the selection curves for the fleets, a reasonable classification of the fishable biomass would be the mature biomass. A corresponding 5\% harvest of this would yield not more than 2.500 tonnes.

## F. References

Björnsson, H., and Sigurdsson, T. 2003. Assessment of golden redfish (Sebastes marinus L.) in Icelandic waters. Scienta Marina 67 (Suppl. 1):301-314.

Bogstad, B., Howell, D., and Åsnes, M. N. 2004. A closed life-cycle model for Northeast A rctic Cod. ICES C.M.2004/K:26, 26 pp.Björnsson and Sigurdsson 2003

Nedreaas, K., 1990. Age determination of Northeast Atlantic Sebastes species. J. Cons. int. Explor. Mer 47, 208-230.

## Annex 8 Quality Handbook

ANNEX:_afwg-ghl-arct
Stock specific documentation of standard assessment procedures used by ICES.

Stock:<br>North-East Arctic Greenland Halibut<br>Working Group: Arctic Fisheries Working Group<br>Date:<br>27-04-09

A. General

## A. 1 Stock definition

Greenland halibut (Reinhardtius hippoglossoides, Walbaum) is distributed in the Arctic and boreal waters in the North Atlantic and in the North Pacific (Fedorov 1971; Godø and Haug 1989; Bowering and Brodie 1995; Bow ering and Nedreaas 2000). In the northeastern Atlantic the distribution is more or less continuous along the continental slope from the Faeroe Islands and Shetland to north of Spitsbergen (Whitehead et al. 1986; Godø and Haug 1989), with the highest concentrations from 500 to 800 m depth between Norway and Bear Island, which is also regarded as the main spawning area (Godø and Haug 1987; Albert et al. 2001b). Peak spawning occurs in December in the main spawning area, but also in nearby localities during summer (Albert et al. 2001b). Atlantic currents transport eggs and larvae northwards and the juveniles are distributed around Svalbard and in the northeastern Barents Sea, to the waters around Franz Josef Land and Novaja Zemlya area (Godø and Haug 1987; Godø and Haug 1989; Albert et al. 2001a). As they grow older they gradually move southwards and eventually alternate between the spawning area and feeding areas in the centralwestern Barents Sea (Nizovtsev, 1989).

The Northeast arctic Greenland halibut stock is a pragmatically defined management unit. The degree of exchange with other stocks is not resolved, but is believed to be low. Potential routes of exchange may be drift of larvae towards Greenland and migration of adults between the Barents Sea and the Iceland-Faeroe Islands area.

## A. 2 Fishery

Before the mid 1960s the fishery for Greenland halibut was mainly a coastal long line fishery off the coasts of eastern Finnmark and Vesterålen in Norway. The annual catch of the coastal fishery was about $3,000 \mathrm{t}$. In recent years this fishery has landed $3,000-6,000 t$ although now gillnets are also used in the fishery. In 1964 dense Greenland halibut concentrations were found by Soviet trawlers in the slope area to the west of the Bear Island (Nizovtsev, 1989). Following the introduction of international trawlers in the fishery in the mid 1960s, the total landings increased to about $80,000 \mathrm{t}$ in the early 1970s. The total Greenland halibut landings decreased steadily to about $20,000 \mathrm{t}$ during the early 1980s. This level was maintained until 1991, when the catch increased sharply to $33,000 \mathrm{t}$. From 1992 total landings varied between 9 000-19 000 t with a peak in 1999.

From 1992 the fishery has been regulated by allowing only the long line and gillnet fisheries by vessels smaller than 28 m to be directed for Greenland halibut. This fish-
ery is also regulated by seasonal closure. Target trawl fishery has been prohibited and trawl catches are limited to bycatch only. From 1992 to autumn 1994 bycatch in each haul was not to exceed $10 \%$ by weight. In autumn 1994 this was changed to $5 \%$ bycatch of Greenland halibut onboard at any time. In autumn 1996 it was changed to 5\% bycatch in each haul, and from January 1999 this percentage was increased to $10 \%$. In August 1999 it was adjusted further to $10 \%$ in each haul but only $5 \%$ of the landed catch. From 2001 the bycatch regulations again was changed to $12 \%$ in each haul and $7 \%$ of the landed catch.

The regulations enforced in 1992 reduced the total landings of Greenland halibut by trawlers from 20,000 to about $6,000 \mathrm{t}$. Since then and until 1998 annual trawler landings have varied between 5,000 and $8,000 \mathrm{t}$ without any clear trend attributable to changes in allowable by catch. How ever, the increase of trawler landings in 1999 to 10 000 t may be attributable partly to the less restrictive bycatch regulations. Landings of Greenland halibut from the directed longline and gillnet fisheries have also increased in recent years to well above the level of $2,500 \mathrm{t}$ set by the Norw egian authorities. This is attributed to the increased difficulties of regulating a fishery that only lasts for a few weeks.

## A. 3 Ecosystem as pects

As investigations show, among the variety of fish, seabirds and marine mammals Greenland halibut were found in the diet of just three species - Greenland shark (Somniosus microcephalus), cod (Gadus morhua morhua) and Greenland halibut itself. Besides, killer whale (Orcinus orca), grey seal (Halichoerus grypus) and narwhal (Monodon monoceros) could be its potential predators. However, the presence of Greenland halibut in the diet of the above species was minor. Predators fed mainly on juvenile Greenland halibut up to $30-40 \mathrm{~cm}$ long.
The mean annual percentage of Greenland halibut in cod diet in 1984-1999 constituted $0,01-0,35 \%$ by weight ( $0,05 \%$ in average) (DOLGOV \& SMIRNOV 2001). Low levels of consumption are related to the distribution pattern of juvenile Greenland halibut as they spend the first years of the life mainly in the outlying areas of their distribution, in the northern Barents Sea, where both adult Greenland halibut and other abundant predator species are virtually absent.
Cannibalism was the highest in 1960's (up to $1,2 \%$ by frequency of occurrence). During the 1980's, in the Greenland halibut stomachs the frequency of occurrence of their own juveniles did not exceed $0,1 \%$. During the 1990's, the portion of their own juveniles (by w eight) was at the level of 0,6-1,3\%.

Food composition of the Greenland halibut in the Barents Sea includes more than 40 prey species (NIZOVTSEV 1989; DOLGOV \& SMIRNOV 2001). Investigations over a wide area of the continental slope up to the Novaya Zemlya show that the main food source of Greenland halibut consists of fish, mostly capelin (Mallotus villosus villosus) and polar cod (Boreogadus saida) followed by cephalopods and shrimp (Pandalus borealis). During the 1990's an important component of the diet was waste products from fisheries for other species (heads, guts etc.). With growth, a decrease in the importance of small food items (shrimp, capelin) in Greenland halibut diet and the increase of a portion of large fish such as cod and haddock (Melanogrammus aeglefinus) were observed.

With the Greenland halibut stock being nearly 100000 tonnes, the total food consumption of the population is estimated to be about 280000 tonnes. The biomass of commercial species consumed (shrimp, capelin, herring, polar cod, cod, haddock,
redfish (Sebastes sp.), long rough dab (Hippoglossoides platessoides) does not exceed 5 000-10 000 tonnes per species (DOLGOV \& SMIRNOV 2001).

The Greenland halibut as a species thus has a negligible effect on the other commercial species in the Barents Sea both as predator and prey.

Greenland halibut occurs over a wide range of depths (from 20 to 2200 m ) and temperatures (from -1.5 to $10^{\circ}$ C) (Boje \& Hareide, 1993; Shuntov, 1965; Nizovtsev, 1989). Young Greenland halibut occur mostly in the northeastern Barents Sea (Spitsbergen archipelago and further east to Franz Josef Land) where the presence adult Greenland halibut or other predators appears minimal. Therefore, Greenland halibut mortality after settling in the area is low and stable and driven mainly by envionmental factors.

## B. Data

## B. 1 Commercial catch

Norwegian commercial catch in tonnes by quarter, area and gear are derived from the sales notes statistics of the Directorate of Fisheries. Data from about 20 sub areas are aggregated on 6 main areas for the gears gill net, long line, bottom trawl and shrimp trawl. For bottom trawl the quarterly area distribution of the catches is adjusted by logbook data from The Directorate of Fisheries and the total bottom trawl catch by quarter and area is adjusted so that the total annual catch for all gears is the same as the official total catch reported to ICES. No discards are reported or accounted for in the catch statistics.

Russian catch based on daily reports from the vessels are combined in the statistics of the All-Russian Research Institute of Fisheries and Oceanography (VNIRO, Moscow). Data are provided separately by ICES areas and gears.

The sampling strategy is to have age-length samples from all major gears in each area and quarter. There are at present no defined criteria on how to allocate samples of catch numbers, mean length and mean weight at age to unsampled catches, but the following general process has been applied: First look for samples from a neighbouring area if the fishery extends to this area in the same quarter. If there are no samples available in neighbouring areas, search for samples from other gears with the most similar selectivity in the same area or in neighbouring areas. The last option is to search in neighbouring quarters, first from the same gear in the same area, and then from neighbouring areas and similar gears. ALKs from research surveys (shrimp trawl) are also used to fill gaps in age sampling data.

Norway and Russia, on average, have accounted for about 90-95\% of the Greenland halibut landings during more recent years. Data on catch in tonnes from other countries are either taken from ICES official statistics (by ICES area) or from reports to Norwegian authorities. A few countries also supply some additional data. The text table below indicates the type of data provided by country:

|  | KIND OF DATA |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Country | Caton (catch in weight) | Canum (catch at age in numbers) | Weca (weight at age in the catch) | Matprop (proportion mature by age) | Length composition in catch |
| Norway | $x$ | $x$ | $x$ |  | $x$ |
| Russia | $x$ | $x$ | $x$ | $x$ | $x$ |
| Germany | $x$ |  |  |  |  |
| United | $x$ |  |  |  |  |
| Kingdom | $x$ |  |  |  |  |
| France ${ }^{1}$ | $x$ |  |  |  |  |
| Spain ${ }^{1}$ | $x$ |  |  |  |  |
| Portugal ${ }^{1}$ | $x$ |  |  |  |  |
| Ireland ${ }^{1}$ | $x$ |  |  |  |  |
| Greenland ${ }^{\text {l }}$ | $x$ |  |  |  |  |
| Faroe Islands ${ }^{1}$ | $x$ |  |  |  |  |
| Iceland ${ }^{\text {l }}$ | $x$ |  |  |  |  |
| Poland ${ }^{1}$ | $x$ |  |  |  |  |

${ }^{1}$ As reported to Norwegian authorities
The Norw egian and Russian input files are Excel spreadsheet files before aggregation to international data. The data are archived in the national laboratories and with the Norw egian stock co-or dinator.

The national data have been aggregated with international data on Excel spreadsheet files. The Russian and Norwegian catch-at-age data based on national landings, length composition of catches, age-length-keys (ALK) and weight at age data. Catches from the other countries were assumed to have the same age composition and weight at age as the Norwegian landings. From 2006 Norway stopped to determine the age using the traditional method. Since than the common catch-at-age files constructed on the base of the Russian ALKand weight at age data.

The Excel spreadsheet files used for age distribution, adjustments and aggregations are held by the Norwegian stock co-ordinator and for the current and previous year in the ICES computer system under w:lacfmlafwglyearlpersonallname (of stock coordinator).

The result files (FAD data) can be found at ICES and with the stock co-ordinator, either in the IFAP system as SAS datasets or as ASCII files on the Lowestoft format, under w: lacom $\operatorname{afw}$ glyearldatalghl_arct.

## B. 2 Biological

For 1964-1969, separate weight at age data are used for the Norwegian and the Russian catches. Both data sets are mean values for the period and are combined as a weighted average for each year. A constant set of weight-at-age data is used for the total catches in 1970-1978. For subsequent years annual estimates are used. The mean weight at age in the catch is calculated as a weighted average of the weight in the catch from Norway and Russia. The weight at age in the stock is set equal to the weight at age in the catch for all years.

A fixed natural mortality of 0.15 is used both in the assessment and the forecast.
Both the proportion of natural mortality before spawning (Mprop) and the proportion of fishing mortality before spawning (Fprop) are set to 0 .

Annual ogives based on sexes combined using Russian survey data are given for the years 1984-1990 and 1992-last data year. An average ogive derived from 1984-1987 is used for 1964-1983. For 1984 to the last data year a three-year running average is used.

## B. 3 Surveys

The results from the following research vessel survey series are evaluated by the Working Group:

1. Norw egian bottom trawl survey in August in the Barents Sea and Svalbard from 1984 in fishing depths of less than 100 m and down to 500 m . (Table E1 and E2).
2. Norw egian Greenland halibut surveys in August from 1994. The surveys cover the continental slope from 68 to $80^{\circ} \mathrm{N}$, in depths of $400-1500 \mathrm{~m}$ north of $70^{\circ} 30^{\prime} \mathrm{N}$, and $400-1000 \mathrm{~m}$ south of this latitude. This series has in 2000 been revised to also include depths between 400 - 500 m in all years (Table E3).
3. Norwegian bottom trawl surveys east and north of Svalbard in autumn from 1996 (Table E4).
4. The Norwegian Combined Survey index Table E5, combination of the results from Tables E1-E4.
5. Russian bottom trawl surveys in the Barents Sea from 1984 in fishing depths of $100-900 \mathrm{~m}$. This series has been revised substantially since the 1998 assessment in order to make the years more comparable with respect to area coverage and gear type (Table E6).
6. Spanish bottom trawl survey in the slope of Svalbard area in October, ICES Division Ilb: from 1997 (Table E7).
7. Norwegian (from 2000 Joint) Barents Sea bottom trawl survey (winter) from 1989 in fishing depths of less than 100 m and down to 500 m . In order to utilise the last year values in the VPA calibration, this series was adjusted back by one year and one age group to reflect sampling as if it occurred in the autumn of the previous year (Table E8).
8. International pelagic 0-group surveys from 1970. (Table 1.1).

Over the last several years the Working Group has been concerned about trends in catchability within individual surveys used for tuning of the XSA. The trends were seen for younger ages of year classes in the late 80's and early 90's that were initially estimated to be very low in abundance. With increasing age these year classes were estimated to be much closer to the mean abundance. In previous meetings the Working Group therefore increased the lower age used in tuning to five years in order to reduce the problem. This only partly resolved the problem though, and in all subsequent assessments estimated recruitment of the last 2-3 years has increased from one year to the next.

The Norw egian bottom trawl survey in the Barents Sea and Svalbard catch Greenland halibut mainly in the range of ages $1-8$, although in most years age 1 is poorly represented and all age group younger than five years are not considered to be well represented in this survey due to the limited depth range covered. The relative strength of the year classes varies considerably with age. In more recent years there has been low but somewhat better representation of young fish in this survey.

The Norwegian juvenile Greenland halibut survey north and east of Svalbard were started in 1996 and from 2000 this survey is conducted as a joint survey between Norway and Russia. As a result it is expected that the area coverage will improve, better representing the distribution of juveniles and will provide a more comparable time series. Only the Norwegian part of these northern surveys is currently included in the Norw egian Combined Survey index (see below) . In future, when the extended coverage in the Russian zone has been repeated for at least five years the Working Group will consider revising the combined index.
The Norw egian Greenland halibut survey along the deep continental slope south and west of Spitsbergen began in 1994. Although Greenland halibut older than 15 years are caught, few fish are represented in the catch over age 12 or less than age 5 (Table E4). Most of the abundance indices are dominated by ages 5-8.
Most of the surveys considered by the Working Group in 2002 cover either the adult population in the slope area or juvenile distribution in northern areas. The problem of underestimation of recruitment in the last few years included in the analyses has been attributed to shortcomings in survey coverage. The Working Group at previous meetings has noted the need for annual surveys that sample most of the population within a short period of time. Prior to the 2002 WG meeting effort was therefore made to combine some of these surveys into a new total index. The new index is termed the Norwegian Combined Survey Index and is established back to 1996, the first year with survey coverage northeast of Svalbard. It includes bottom trawls from the Norwegian bottom trawl survey in August in the Barents Sea and Svalbard (Tables E1 and E2), the Norwegian Greenland halibut survey in August along the continental slope (Table E3), and the Norwegian bottom trawl survey in AugustSeptember north and east of Svalbard (Table E4). Prior to the meeting in 2003 work was done to evaluate the combination of these survey series into one index and this was reported in Working Document 5 to the Working Group. Based on these results it was decided to use this combined index in this years assessment.
The Norwegian Combined Survey Index (Table E5) indicates a significant increase in the total stock during the last three years and a stock size in 2002, nearly $40 \%$ above last years index. However, there is no clear year class pattern in the data and some ages are consistently underestimated relative to adjacent age groups (e.g. age 9 and partly age 4). The highest indices were observed for age seven, with exception of the two last years when age 1 was most abundant. That indicates that the catchability of younger ages (i.e. those primarily from northern surveys) are not comparable with the older ones (i.e. those primarily from the slope). This is probably a result of pooling different surveys using different gears. These weaknesses reduce the applicability of the combined surveys, and the Working Group advises that further work be done to improve the combined index in the future.

The Russian Barents Sea bottom trawl survey, which extends back to 1984 catch fish mainly in the range of $4-10$ years old. The relative abundance of the year classes against age is similar to the surveys above. This survey covers the Barents Sea including the continental slope of the Norwegian Sea. Total abundance indices from this survey show trend to grow since 1996.

The Spanish bottom trawl surveys along the continental slope north of $73^{\circ} 30^{\prime} \mathrm{N}$ from 1997 (Table E7) differ from the other survey series indicating reduced abundance in this area since 1999.

The Norwegian bottom trawl survey during winter in the Barents Sea catch Greenland halibut older than 12 years, but are not particularly effective in catching
fish older than 7 years. This is likely due to the limited depth distribution of the survey area. Nevertheless, the survey appears very effective at catching Greenland halibut up to age 6 . The relative abundance of the year classes against age is comparable with the survey above.
The strengths of the Greenland halibut year classes of 1970-1997 from the International pelagic 0-group surveys in the Barents Sea are shown in Table 1.1. The results are highly variable over the time period. However, most of the 1970's and 1980's year classes are represented in reasonably high numbers. In recent years the 1988-1992 and the 1996 year classes have been well below the long term average. The 1993-1995 and 1997-1999 year classes are closer to the average. Significant increase of 0-group abundance indices with compare to previous years was observed in 2000-2002. Than the increase in 0-group abundance seems to have stopped, and the 2007-2008 indices were very low. It should be noted that the Ecosystem survey is not optimal for surveying 0 -group Greenland halibut.
All in all, the surveys seem to indicate that the catchability of the 1990-1995 year classes increased considerably as the fish becomes five years and older. Based on extremely low catch rates in the surveys, these year classes were considered very poor in previous assessments by the W orking Group, but improved considerably at older ages. The reason for this change in catchability is not clear. However, it is known that important areas for young Greenland halibut may be found north and east of Svalbard (Table E4). (Albert et al. 2001a) showed that the south-western end of the distribution area of age 1 fish was gradually displaced northwards along west Spitsbergen in the period 1989-92 and southwards in the period 1994-1996. These displacements corresponded to changes in hydrography and may be explained by increased migration of the 1990-1995 year classes to areas outside the survey area.
Since 2006, none of the age structured tables of the Norwegian surveys have been updated due to change in age reading procedure.

## B. 4 Commercial CPUE

The restrictive regulations imposed on the trawl fishery after 1991 disrupted the traditional time series of commercial CPUE data. However, an attempt to continue the series was made through a research program using two Norwegian trawlers in a limited commercial fishery (Tables 8.6 and E9). This comprises fishing during two weeks in May-June and October, representing an effort somewhat less than $20 \%$ of the 1991 level. Since 1994 the fishery has been restricted to May-June. This fishery was conducted, as much as possible, in the same way as the commercial fishery in the previous years. The Norw egian CPUE survey was stopped from 2005. This was one of the tuning fleets, but an evaluation of this survey revealed a lot of inconsistencies in the series.

Since 1997 also two Russian trawlers conducted a limited research fishery for Greenland halibut.

The CPUE from the experimental fishery was found, however, to be considerably higher than in the traditional fishery and has exhibited an increasing trend from 1992-1996. After 1996 the Norwegian CPUE series has varied between 1200 and $1650 \mathrm{~kg} / \mathrm{h}$ with the highest value in 2000 (Table E9). The Russian experimental CPUE series shows an increasing trend since 1997, and this series also shows the highest value in 2000.

## B. 5 Other relevant data

None

## C. Historical stock development

Model used: XSA
Software used: IFAP / Low estoft VPA suite
Model Options chosen:
Tapered time weighting applied, power $=3$ over 20 years
Catchability independent of stock size for all ages
Catchability independent of age for ages $>=10$
Survivor estimates shrunk towards the mean $F$ of the final 2 years or the 5 oldest ages
S.E. of the mean to which the estimate are shrunk $=0.500$

Minimum standard error for population estimates derived from each fleet $=0.300$
Prior weighting not applied
Input data types and characteristics:

| Type | Name | Year range | Age range | V ariable from year to year <br> Yes/No |
| :---: | :---: | :---: | :---: | :---: |
| Caton | Catch in tonnes | 1964 - last data year | - (total) | Yes |
| Canum | Catch at age in numbers | 1964 - last data year | 5-15+ | Yes |
| Weca | Weight at age in the commercial catch | 1964 - last data year | $5-15+$ | Yes/No - constant at age from 1964-1978 |
| West | Weight at age of the spawning stock at spawning time. | $1964 \text { - last data }$ year | 5-15+ | Yes/No - assumed to be the same as weight at age in the catch |
| Mprop | Proportion of natural mortality before spawning | 1964 - last data year | $5-15+$ | No - set to 0 for all ages in all years |
| Fprop | Proportion of fishing mortality before spawning | 1964 - last data year | 5-15+ | No - set to 0 for all ages in all years |
| Matprop | Proportion mature at age | 1964 - last data year | $5-15+$ | Yes/No - three year running mean, constant at age from 1964-1983 |
| Natmor | Natural mortality | 1964 - last data year | 5-15+ | No - set to 0.15 for all ages in all years |

Tuning data:

| Type | Name | Year range | Age range |
| :--- | :--- | :--- | :--- |
| Tuning fleet 1 | Norwegian Combined <br> survey index | 1996 - last data year | $5-15+$ |
| Tuning fleet2 | Norwegian <br> experimental CPUE | 1992 - last data year | $5-14$ |
| Tuning fleet3 | Russian trawl survey <br> from 1992 | 1992 - last data year | $5-15+$ |

## D. Short-term projection

Model used: Age structured
Software used: IFAP prediction with management option table and yield per recruit routines

Initial stock size. Taken from the XSA for age 6 and older. The recruitment at age 5 in the last data year is estimated using the mean from 1990 to two years before the last data year following the argument that recruitment at age 5 shows a sharp reduction in the most recent years in the previous assessments, which is not believed to reflect the true recruitment.

Natural mortality: Set to 0.15 for all ages in all years
Maturity: The same ogive as in the assessment is used for all years
F and Mbefore spawning: Set to 0 for all ages in all years
Weight at age in the stock: Average weight at age for the last three years used in the assessment

Weight at age in the catch: Average weight at age for the last three years used in the assessment

Exploitation pattern: Average of the three last years
Intermediate year assumptions: Catch constraint
Stock recruitment model used: Constant recruitment as described earlier
Procedures used for splitting projected catches: Not relevant

## E. Medium-term projections

Not done
F. Long-term projections

Not done

## G. Biological reference points

No limit or precautionary reference points for the fishing mortality or the spawning stock biomass are proposed.

## Other issues

None

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## Issues related to catch data collection and methodological im-provements:-AFWG

| Stock | Data Problem | How to be addressed | By who |
| :---: | :---: | :---: | :---: |
| Stock name | Data problem identification | Description of data problem and recommend solution | Who should take care of the recommended solution and who should be notified on this data issue. |
| Ne A saithe | Reduction in samples north of 67 N from Q3 2009 for gillnet, Danish se ine and line | The sampling should be improved from 2011 onwards | Norway |
| NEA saithe | Lack of purse seine samples be tween 6267N | The sampling should be improved from 2011 onwards | Norway |
| NEA saithe | Lack of useful recruitment indices of 1 year olds |  |  |
| NEA cod | Recruitment indices | Study group for recruitement | PGCCDBS |
| Ne A saithe | In FLR " 0 " values in Canum file is not handled properly | Observe d in WKROUND 2010 for Ne A saithe where age span was expanded from $11+$ to $15+$ and for some years the re are " 0 " values in canum file for somage groups. Bug reported to FLR web site http://bugs.flr-proje ct.org/ | FLR de ve lopment team |
| Afwg stocks | Aiming at changing the basis for its advice from $\mathrm{F}_{\mathrm{pa}}-\mathrm{B}_{\mathrm{pa}}$ to Fmsy, combine d with a trigger spawning biomass (Btriger). <br> How should MSY be analysed and evaluated and be se lecte d compared to excisting values in use.Which Year span and othersettings is re presentative for long term Yield conside rations for each stock? Is it a new targe tpoint or range? | Work shop of MSY-related studies, in be tween afwg meetings? | ACOM |


| Stock | Data Problem | How to be addressed | By who |
| :---: | :---: | :---: | :---: |
| NEA haddock | Systematic differences in weight at age when comparing Russian surveys in late autumnand Norwe gian surveys in winter. Possibly an age-reading problem. | First, the actual diffe rences should be investigated further, e.g. by region, to exclude other possible sources of error Second, age reading comparisons should be intensified to investigate and possibly remedy be tween-rea der bias | IMR, PINRO |
| Sebastes mentella | Norwe gian and Russianage readings are not properly harmonized for mature fish, especially above age 15 | The ICES Workshop on Age Determination of Redfish (WKADR) has re ported this problem to be related to not including the proximal zone of the otolith sections when reading and determining the age. | Frequentexchanges of otoliths between Norway, Russia and others for comparative age readings. Should be re ported to PGCCDBS and AFWG. |
| Sebastes mentella | Not all countries fishing S. mentella in international waters of the Norwegian Sea re port the ir catches to NEAFC and ICES. EU reported catches are not split by individual country, which is problematic for the assessment. Lack of cons istency between daily reports from the sea to NEAFC and later official reports by de le gates to NEAFC. | NEAFC should provide ICES and AFWG with both the daily re ports from the sea and the official re ports to NEAFC by delegates. | PGCCDBS, ACOM, RCM NSEA |
| Sebastes spp. | Reduction in samples from the commercial fisheries for S. mentella and $S$. marinus | The sampling should be improved from 2010 onwards | RCM NSEA, Russia, Norway, ACOM |

## Annex 10 Evaluation of Rebuilding plan for coastal cod

Request from the Royal Norwegian Ministry of Fisheries and Coastal Affairs

## Rebuilding plan for Norwegian coastal cod

The Norwegian coastal cod north of $62^{\circ} \mathrm{N}$ is recognized as a stock complex. Genetic studies indicate that some of the spawning components along the coast could be local stocks - more or less isolated from coastal cod in adjacent areas. Subsequently, the coastal cod management faces two major challenges - those being, first, to keep the total stock complex at a productive level and second, to give protection to potentially vulnerable local stock components. Both of these challenges are addressed in this draft rebuilding plan for coastal cod. Moreover, the knowledge regar ding local stocks should be more specified, due to the fact that the scientific advice provided has become increasingly more specific on these matters.

The catch at age analysis (xsa) prepared for this stock has been considered uncertain. This is mainly due to the shortage of information from the recreational and tourist fisheries respectively. Similarly, the division of Northeast Arctic cod and coastal cod in the catches in northern areas has also been deemed uncertain. As a result of this uncertainty, the analyses of annual updates of spawning stock and fish mortality has not been particularly useful when seeking to define reference values for a rebuilding plan. Nevertheless, the coastal survey time series from 1995 onwards could be applied to define a sufficient rebuilding target. Similarly, mortality signals from sampling data could be used to monitor changes in fishing mortality.

## Rebuilding plan

The overarching aim is to rebuild the stock complex to full reproductive capacity, as well as to give sufficient protection to local stock components. Until a biologically founded rebuilding target is defined, the stock complex will only be regarded as restored when the survey index of spawning stock in tw o successive years is observed to be above 60000 tons ${ }^{1}$. Importantly, this rebuilding target will be redefined on the basis of relevant scientific information. Such information could, for instance, include a reliable stock assessment, as well as an estimate of the spawning stock corresponding to full reproductive capacity.

Given that the survey index for ssb does not increase, the regulations will aim to reduce $F^{2}$ by at least 15 per cent annually compared to the $F$ estimated for 2009. If, however, the latest survey index of ssb is higher than the preceding one - or if the estimated F for the latest catch year is less than 0.1 - the regulations will be unchanged.

Special regulatory measures for local stock components will be view ed in the context of scientific advice. A system with stricter regulations inside fjords than outside fjords is currently in operation, and this particular system is likely to be continued in the future.

The management regime employed is aiming for improved ecosystem monitoring in order to understand and possibly enhance the survival of coastal cod. Potential predators are - among others - cormorants, seals and saithe.

[^13]When the rebuilding target is reached, a thorough management plan is essential. In this regard, the aim will be to keep full reproductive capacity and high long-term yield.
ICES is asked to evaluate whether the above rebuilding plan is consistent with the precautionary approach. If this is not the case, or if the basis for evaluation is unsatisfactory, further advice for modifications or alternative plans is requested.

## Comments on regulatory measures

At present, there are several regulatory measures employed. Importantly, the commercial catches of coastal cod are currently taken by vessels that have quotas of Northeast Arctic cod, whereby a small quantity has been added to their quota in addition to the expected "by-catch" of coastal cod. Second, the core regulation strategy used to reduce catches of coastal cod has been to restrict parts of the fleet to areas and seasons were they are most likely to catch Northeast Arctic cod. Third, since 2004 only vessels less than 15 meters have been allowed to fish within the fords, as defined by ford-lines. Moreover, only vessels less than 21 meters have been allowed to fish between the base lines and the ford lines. Fourth, long-liners fishing with automatic baiting have to fish outside 4 nautical miles ( nm ), whereas trawlers have to fish outside $12 \mathrm{~nm}{ }^{3}$, Similarly, vessels fishing with Danish seine have to fish outside the fjord lines. Fifth, two coastal cod spawning areas have - in the spawning season been closed for fishing, except for fishing with hand lines. Finally, some restrictions to the recreational fishery have also been introduced.

All the aforementioned regulation measures can potentially be employed to further restrict catches of coastal cod. In addition to these measures, a principle of increased mesh size can be introduced in coastal areas. This will improve the likeliness of survival to age of spawning, and further, the survival of second time spawning.

Due to the complexity of these fisheries and the variable mixing between Northeast Arctic cod and coastal cod, the exact annual effect of gradually increased regulations has proved difficult to predict. The accumulated effects over several years should, however, be expected to be in line with the above rebuilding plan.

## Approach for the evaluation

The essence of the Rebuilding plan:

> -Reduce F annually by $15 \%$ relative to $F 2009$.
> -If the latest survey shows an increase, or if latest $F$ is estimated to be $<0.10$, the regulations shall remain unchanged.

Rebuilding target $=$ average survey SSB in the period 1995-1998.
Assessment in spring in year Y , using survey and catches up to year $\mathrm{Y}-1$
On this basis give management advice for year $\mathrm{Y}+1$ (conditional to the survey result in fall in year y)

This reduction rule means that, conditional to the survey results, the following year will either require further action, or status quo. The starting value of F 2009 will determine the number of action years required to reach the lower limit. As an example:
${ }^{3} 6 \mathrm{~nm}$ in some areas

F2009 $=0.3$ will represent action steps of 0.045 , which means 5 steps for theoretically getting below the lower limit. The total time span for reaching the lower F-limit will then be 5 plus the number of years with increased survey estimates. This is illustrated in the scheme below, where no errors are assumed for $F$, the starting point is 0.3 and the last step is adjusted to hit 0.1 . In reality there will also be some additional action years in cases when the real F is below threshold, while estimated above, and at the same time the survey does not increase.

Coastal cod: Simulation of recovery plan
Individual trajectaries of $F$ - no dbs. or implementationerror

[The reduction rule is here taken as fixed steps. The wording " $15 \%$ annually" could possibly be interpreted as $15 \%$ relative to the latest reduction. This means gradually decreasing steps, which prolongs the whole process and, after some years, causes "micro steps" that would be unrealistic compared to the precision both for the regulation measures and the stock assessment.]

The request is rather open both in terms of the basis for rebuilding target ("full reproduction potential", reference to historic survey results, or reference to historic SSB from analytical assessment). There is thus a need for considering candidates for "full reproductive potential" (analogous to Blim), and considering improvements of the assessment. The HCS software (see details under "simulation model" p621) was considered useful both for examining candidate reference points and for simulating the plan.

The Norway-Russia annual quota agreements specify an expected catch of Norwegian coastal cod (NCC). This is a part of the total quota balance for cod. In the Norwegian regulations there is no specific quota for fishing coastal cod, but a total quota for cod, which typically is about 10 times larger than the expected "by-catch" of coastal cod. Thus the fishery for coastal cod is not directly regulated by quotas. Regulations introduced for reducing catches of NCC are closures of areas and seasons, and restrictions on vessel size and gear types/mesh size. These regulations shift the effort from fishing NCC to fishing NEAC or other species. For NCC this is thereby a type of effort regulation, which is an appropriate approach for regulating F. Originally the HCS simulation soft ware was designed to evaluate TAC-regulations where the intended F were translated to TAC, and the realized F was the result of the realized catch under the given TAC. Several simulations were done with this version, in case more direct TAC-regulations are part of the future plan. The main simulations were made with a modified, pure F-based version of HCS to better reflect the purpose of gradual F-reductions.

The complexity of these fisheries leads to a rather complex relation betw een the regulation and the effective effort "hitting" NCC. The implementation of the planned Freductions will therefore be rather uncertain, and the simulations need to assume a rather large implementation error.
The F-reduction is conditional to the survey result. The consequence is that the uncertainty of the survey will contribute to slowing down the F-reduction rate, at least when the true stock is stable or declining. The simulations are therefore set up to examine different assumed survey uncertainties.
The xsa-assessment for NCC has shown rather serious retrospective problems and recent stock assessments have mainly been based on inspections of survey trends and signals in the catch data. In such a situation it is, therefore, convenient to consider a rebuilding target relating to future survey results. Since the plan is expressed in terms of F-reductions it is essential to be able to estimate recent Fs with a reasonable accura$c y$, thereby monitoring the results of future regulations following this plan. The retrospective XSA-analysis (made according to the existing Stock Annex) shows a quite biased F for the unconverged period (Figure 1). Trends of Z and F from "Surba" analysis has been tried at recent AFWGs, but currently warnings are given that the program should not be trusted. The next chapter describes how alternative methods can give reasonable estimates for the current F . Such estimates can be used further to estimate current stock size from the catch at age (defining terminal F in a traditional vpa), thereby giving an SSB that can be compared to a historic estimate of SSB. AFWG has this year used this approach to obtain an improved analytical assessment. This may be further improved before next benchmark assessment, and hopefully provide the basis for an alternative reference for the rebuilding target, in case the survey would be unsuccessful or cancelled.


Figure 1. Retrospective Fin xsa with the input and settings used by AFWG09.

## Alternative methods for estimation of $F$

## Survey based mortality estimates

A simple approach is to use the survey data, taking the decline in survey index ( U ) of each cohort from one year ( $U_{a y}$ ) to the next ( $\mathrm{U}_{\mathrm{a}+1, \mathrm{y}+1}$ ). This would contain all mortality,
and in addition it will contain the age dependence in catchability. If both catchability and natural mortality is considered stable between years, those factors will only influence the scaling of the "survey mortality" while the trends observed would be driven by F. The coastal cod survey takes place late in the year, and it is reasonable to define the survey mortality for age a in year y as:
$Z_{a, y}=-\log \left(U_{a+1, y+1} / U_{a y}\right)$.
These age specific values can be further averaged within years for those ages where the survey is considered to bebest. The resulting survey Zs by age and average across ages are shown in Table 1 and Figure2.

Table 1. Survey Z at age and averaged for ages $4-9$, xsa values and F predicted from the surve y Z . $R$ is the correlation with vpa Fs for the corresponding age groups over the period 96-05.

|  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | av4-9 | $\mathrm{F}(4-7) \mathrm{vpa}$ | pred F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 |  |  |  |  |  |  |  |  |  | 0.306 |  |
| 1996 | 0.502 | -0.122 | 0.096 | 0.223 | 0.867 | 1.100 | 1.437 | 1.566 | 0.881 | 0.380 | 0.381 |
| 1997 | -2.372 | -0.509 | 0.275 | 0.000 | 0.165 | 0.551 | 0.741 | 3.367 | 0.850 | 0.405 | 0.379 |
| 1998 | 0.810 | 0.227 | 0.781 | 0.849 | 1.373 | 1.906 | 2.523 | 2.193 | 1.604 | 0.441 | 0.443 |
| 1999 | 0.245 | 0.115 | 0.393 | 0.610 | 0.888 | 1.306 | 1.875 | 1.034 | 1.018 | 0.441 | 0.393 |
| 2000 | -0.521 | -0.026 | 0.305 | 0.237 | 0.278 | 0.805 | 0.555 | 1.046 | 0.538 | 0.379 | 0.352 |
| 2001 | 0.354 | 0.366 | 0.427 | 0.636 | 1.135 | 1.295 | 1.031 | 0.949 | 0.912 | 0.322 | 0.384 |
| 2002 | 1.029 | 0.495 | 0.481 | 0.740 | 0.606 | 0.557 | 1.297 | 2.821 | 1.084 | 0.393 | 0.398 |
| 2003 | -0.479 | -0.170 | 0.056 | 0.572 | 0.415 | 0.818 | 0.915 | 0.114 | 0.482 | 0.378 | 0.347 |
| 2004 | -0.530 | -0.544 | -0.198 | 0.341 | 0.363 | 1.115 | 0.937 | 1.793 | 0.725 | 0.334 | 0.368 |
| 2005 | 0.557 | 0.005 | 0.145 | 0.295 | 0.484 | 0.549 | 0.350 | 0.306 | 0.355 | 0.309 | 0.336 |
| 2006 | -0.560 | -0.787 | -0.071 | -0.082 | 0.249 | -0.063 | 1.032 | 0.880 | 0.324 | 0.370 | 0.334 |
| 2007 | -0.537 | -0.480 | -0.310 | 0.149 | 0.350 | 0.552 | 0.868 | 0.706 | 0.386 | 0.367 | 0.339 |
| 2008 | 0.010 | 0.288 | 0.354 | 0.966 | 0.983 | 1.099 | 1.276 | 0.868 | 0.925 | 0.457 | 0.385 |
| 2009 | 0.033 | -0.222 | -0.470 | 0.121 | 0.271 | -0.032 | 0.027 | -0.095 | -0.030 |  | 0.304 |
| R | -0.453 | -0.181 | 0.641 | 0.188 | -0.087 | 0.278 | 0.666 | 0.327 | 0.665 |  |  |



Figure 2. Survey $Z$ averaged over ages 4-9.
When using such mortalities to monitor the effect of new regulations it would be a great advantage to normalize the level of the values so that they can be compared to F values used in the analytical assessment and in the model used for simulating the rebuilding plan. The survey mortalities are therefore regressed against the Fs in the converged part of the xsa. Figure 3 shows the relationship. This relationship is further used to convert the survey mortalities to "predicted xsa-Fs". The time trend of such predictions is shown in Figure 6. A weakness of this data series is the small range of Fs experienced over the observation period.


Figure 3. The relation between survey $Z$ and xsa $F$ for the overlapping part of the converged period (1996-2005).

## Catch based mortality estimates

Similar calculation of $Z$ can be made from the catch matrix. In this case $\log \left(C_{a, y} / C_{a+1, y+1}\right.$ ) is functionally related to $\mathrm{F}_{\mathrm{y}}$ and $\mathrm{F}_{\mathrm{y}+1}$, but somewhat more related to $\mathrm{F}_{\mathrm{y}}$ than to $\mathrm{F}_{\mathrm{y}+1}$. It is therefore reasonable to define
$\mathrm{Z}_{\mathrm{a}, \mathrm{y}}=\log \left(\mathrm{C}_{\mathrm{a}, \mathrm{y}} / \mathrm{Ca}_{\mathrm{a}+1, \mathrm{y}+1}\right)$.
As for the survey mortalities, these catch based year to year mortalities can be used for predicting "vpa-Fs". Results for the coastal cod commercial catch data are shown in Table 2 and Figures 4, 5 and 6. Figure 7 and Table 3 show the results of similar analysis with the new data on recreational catches added.
Both the survey mortality and the catch based mortality show a fairly good ability to follow the historic variations in F seen in the vpa. It is not surprising that the catch based one follow the variation in vpa, but it is interesting to note that this simple mortality estimate appears more precise than the Fs that have been estimated in the unconverged part of the vpa. One disadvantage of the catch method is that it only gives values up to the second last catch year. A value for the latest year could be obtained by scaling the second last year by fishing effort or simply scaling by catch in tonnes.

Table 2. Age specific mortalities estimated from commercial catch at age data. R is the correlation with vpa Fs for the corresponding age groups over the period 96-05.

| 3 | 4 | 5 | 6 | 7 | 8 | 9 | av4-7 | $\mathrm{F}(4-7)$ xsa | Catch pred F |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1984 | -0.751 | -0.224 | 0.027 | 0.672 | 1.952 | 1.608 | 1.132 | 0.607 | 0.622 | 0.499 |
| 1985 | -0.478 | -0.190 | 0.414 | 0.605 | 0.881 | 0.723 | 0.062 | 0.428 | 0.528 | 0.430 |
| 1986 | -0.699 | 0.018 | 0.699 | 0.775 | 1.233 | 1.401 | 0.831 | 0.681 | 0.581 | 0.528 |
| 1987 | -1.269 | 0.244 | 0.629 | -0.057 | 0.595 | 0.819 | -0.062 | 0.353 | 0.492 | 0.401 |
| 1988 | 0.292 | 0.374 | 0.216 | 0.890 | 1.597 | 1.823 | 1.754 | 0.770 | 0.620 | 0.562 |
| 1989 | -0.062 | 0.075 | 0.711 | 0.408 | 1.258 | 1.907 | 1.173 | 0.613 | 0.376 | 0.501 |
| 1990 | -1.426 | -0.926 | -0.342 | -0.139 | 1.376 | 1.767 | 0.405 | -0.008 | 0.184 | 0.261 |
| 1991 | -1.650 | -1.114 | 0.040 | -0.038 | 0.368 | 0.042 | -0.748 | -0.186 | 0.171 | 0.192 |
| 1992 | -0.942 | 0.299 | 0.499 | -0.026 | 0.039 | 0.359 | 0.193 | 0.203 | 0.235 | 0.343 |
| 1993 | -1.523 | -0.925 | -0.052 | -0.218 | 0.708 | 0.879 | -0.241 | -0.122 | 0.236 | 0.217 |
| 1994 | -1.409 | -1.120 | -0.254 | -0.282 | 0.830 | 1.000 | 0.585 | -0.206 | 0.237 | 0.184 |
| 1995 | -1.018 | -0.743 | 0.110 | 0.288 | 0.451 | 0.483 | 0.025 | 0.027 | 0.306 | 0.274 |
| 1996 | -0.282 | 0.094 | 0.683 | 0.166 | 0.228 | 0.775 | 0.272 | 0.293 | 0.380 | 0.377 |
| 1997 | -0.331 | -0.245 | -0.006 | 0.477 | 0.768 | 1.223 | 1.028 | 0.248 | 0.405 | 0.360 |
| 1998 | 0.184 | 0.116 | 0.372 | 0.349 | 0.722 | 0.858 | 1.114 | 0.390 | 0.441 | 0.415 |
| 1999 | -1.019 | -0.484 | 0.468 | 0.921 | 1.592 | 1.739 | 1.628 | 0.624 | 0.441 | 0.506 |
| 2000 | -0.638 | 0.209 | 0.645 | 0.640 | 0.711 | 0.556 | 0.104 | 0.551 | 0.379 | 0.477 |
| 2001 | -1.529 | -0.429 | 0.081 | 0.262 | 0.432 | 0.511 | -0.701 | 0.086 | 0.322 | 0.298 |
| 2002 | -0.852 | -0.072 | 0.268 | 0.691 | 0.945 | 1.266 | 1.189 | 0.458 | 0.393 | 0.441 |
| 2003 | 0.181 | 0.201 | 0.320 | 0.402 | 0.814 | 0.973 | -0.023 | 0.434 | 0.378 | 0.432 |
| 2004 | -1.446 | -0.682 | 0.176 | 0.490 | 1.035 | 1.175 | 0.458 | 0.255 | 0.334 | 0.363 |
| 2005 | -1.251 | -0.266 | 0.076 | 0.142 | 0.508 | 0.453 | -0.175 | 0.115 | 0.309 | 0.309 |
| 2006 | -1.489 | -0.109 | 0.303 | 0.721 | 1.159 | 0.964 | 0.509 | 0.519 | 0.370 | 0.465 |
| 2007 | -1.027 | -0.158 | 0.404 | 0.373 | 0.627 | 0.598 | 0.409 | 0.312 | 0.367 | 0.385 |
| 2008 |  |  |  |  |  |  |  |  | 0.457 |  |
| R | 0.732 | 0.740 | 0.717 | 0.729 | 0.879 | 0.786 | 0.810 | 0.846 |  |  |



Figure 4. Catch based $Z$ averaged over ages 4-7.


Figure 5. The relation between catch based $Z$ and usa $F$ for the overlapping part of the converged period (1984-2005).


Figure 6. Comparison of F-estimates from survey, F-estimates from commercial catches and the converged part of the usa based on commercial catch. 2009 data are included for the survey, but not for the catches.


Figure 7. Comparison of F-estimates from survey, F-estimates from total catches and the converged part of the xsa based on total catches. 2009 data are included for the survey, and for the catches.

Table 3. Age specific mortalities estimated from catch at age including recreational catches. $R$ is the correlation with vpa Fs for the corresponding age groups over the period 96-05.

| 3 | 4 | 5 | 6 | 7 | 8 | 9 | av4-7 | F(4-7) xsa | Catch pred F |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1984 | -0.625 | -0.141 | 0.146 | 0.783 | 1.896 | 1.544 | 1.283 | 0.671 | 0.618 | 0.523 |
| 1985 | -0.384 | -0.082 | 0.416 | 0.586 | 0.772 | 0.552 | 0.230 | 0.423 | 0.524 | 0.416 |
| 1986 | -0.541 | 0.048 | 0.733 | 0.805 | 1.247 | 1.466 | 1.172 | 0.708 | 0.590 | 0.539 |
| 1987 | -0.522 | 0.406 | 0.517 | -0.031 | 0.661 | 0.738 | 0.053 | 0.388 | 0.507 | 0.401 |
| 1988 | 0.132 | 0.191 | 0.211 | 0.990 | 1.600 | 1.773 | 1.927 | 0.748 | 0.634 | 0.556 |
| 1989 | -0.225 | 0.371 | 0.055 | 0.526 | 1.342 | 2.014 | 1.308 | 0.574 | 0.383 | 0.481 |
| 1990 | -0.881 | -0.584 | -0.287 | 0.714 | 1.386 | 1.427 | 0.405 | 0.307 | 0.237 | 0.365 |
| 1991 | -0.511 | -0.285 | 0.207 | 0.086 | 0.075 | 0.059 | -0.318 | 0.021 | 0.194 | 0.242 |
| 1992 | -0.431 | 0.557 | 0.516 | -0.040 | 0.171 | 0.561 | 0.289 | 0.301 | 0.249 | 0.363 |
| 1993 | -0.731 | -0.516 | -0.174 | -0.127 | 0.681 | 0.833 | 0.221 | -0.034 | 0.236 | 0.218 |
| 1994 | -0.905 | -0.664 | -0.352 | -0.076 | 0.763 | 0.951 | 0.733 | -0.082 | 0.240 | 0.197 |
| 1995 | -0.890 | -0.566 | 0.077 | 0.254 | 0.465 | 0.507 | 0.305 | 0.058 | 0.302 | 0.257 |
| 1996 | -0.341 | -0.011 | 0.542 | 0.182 | 0.324 | 0.846 | 0.498 | 0.259 | 0.362 | 0.345 |
| 1997 | -0.540 | -0.246 | 0.056 | 0.500 | 0.866 | 1.179 | 1.157 | 0.294 | 0.401 | 0.360 |
| 1998 | 0.043 | 0.137 | 0.369 | 0.377 | 0.723 | 0.773 | 1.346 | 0.402 | 0.413 | 0.406 |
| 1999 | -0.991 | -0.339 | 0.512 | 0.936 | 1.464 | 1.371 | 1.879 | 0.643 | 0.411 | 0.511 |
| 2000 | -0.523 | 0.181 | 0.586 | 0.487 | 0.671 | 0.344 | 0.706 | 0.481 | 0.352 | 0.441 |
| 2001 | -1.316 | -0.410 | 0.088 | 0.328 | 0.602 | 0.412 | -0.161 | 0.152 | 0.325 | 0.298 |
| 2002 | -0.840 | -0.079 | 0.231 | 0.702 | 0.930 | 0.761 | 1.599 | 0.446 | 0.367 | 0.426 |
| 2003 | 0.029 | 0.045 | 0.159 | 0.303 | 0.635 | 0.775 | 0.794 | 0.285 | 0.336 | 0.356 |
| 2004 | -1.251 | -0.647 | 0.142 | 0.467 | 0.886 | 0.977 | 0.998 | 0.212 | 0.333 | 0.324 |
| 2005 | -1.279 | -0.269 | 0.003 | 0.209 | 0.618 | 0.726 | 0.486 | 0.140 | 0.305 | 0.293 |
| 2006 | -1.364 | -0.049 | 0.321 | 0.681 | 0.951 | 0.870 | 0.838 | 0.476 | 0.336 | 0.439 |
| 2007 | -0.929 | -0.138 | 0.399 | 0.387 | 0.659 | 0.721 | 0.888 | 0.327 | 0.302 | 0.374 |
| 2008 | -0.424 | 0.083 | 0.094 | 0.123 | 0.425 | 0.984 | 0.235 | 0.181 | 0.292 | 0.311 |
| R | 0.635 | 0.583 | 0.832 | 0.778 | 0.791 | 0.858 | 0.683 | 0.819 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |

## Estimates of F2009 and future F-reductions

The current stock assessment has used these alternative estimates. For the time series based on only the commercial catch F2009 is estimated to 0.37 . For the time series including recreational fisheries $\mathrm{F} 2009=0.31$. The corresponding Fs to aim at in future "Action years" will then be as follows:

| Assessment <br> basis | F2009 | Action yr1 | Action yr2 | Actionyr3 | Action yr4 | Action yr5 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| commercial | 0.37 | 0.31 | 0.26 | 0.20 | 0.15 | 0.10 |
| Total catch | 0.31 | 0.26 | 0.22 | 0.17 | 0.12 | 0.10 |

## Simulations

## Simulation model

Simulations were done with a modification of the HCS10 software. The simulation program HCS has been used to explore and evaluate several harvest rules in the past (e.g. mackerel and Blue whiting). It has developed gradually; the most recent update from 2010 is termed HCS10. This program was used recently to explore candidate FMSY values for a range of stocks (Skagen, WD to WKFRAME) and by HAWG for the same purpose for most herring stock covered by that group (ICES 2010 a and b).

The program has not been formally published but a full description of the program, as well as the source code, are available from the author (dankert.skagen@imr no).

The program was modified to simulate a harvest rule that is only guided by a noisy survey estimate of SSB (instead of an assessment result). The program is a stochastic medium term simulation, with 1000 replicas run over 20 years for each set of management options. It has an age structured true population that is projected forwards with randomly drawn recruitments, and true catches emerging from implementing (with noise) a TAC set according to a harvest rule. The harvest rule sets a TACbased on noisy observations of the true stock. In the present version, the observations are SSBs derived by adding noise to the true stock numbers. The rule specifies an F-value that is reduced if the observed SSB is reduced from one year to the next, but with a lower bound of 0.1 . An alternatively option available for simulation is that the F is reduced irrespective of the observed SSB until it reaches 0.1 . The reduction is by a fixed step size, which is a given percentage of the $F$ in the intermediate year at the start of the simulation. The implementation of the rule is by deriving the catch corresponding to the decided F applied to the true stock, adding random noise to the catch numbers at age. This realized catch at age is normalized so that the total catch in tonnes is unchanged. On top of that, the numbers are expanded with a random year multiplier. These implemented catches are removed from the true population.
In the present study, recruitments were drawn according to a hockey stick stockrecruit function with lognormal noise. Weights and maturities at age had fixed values. The assumed selection at age also had fixed values; the realized selection is influenced by the age factor in the implementation error. The F for year Y was reduced if the observed (survey) SSB in year Y-1 was lower than in year Y-2. Simulations started at 1. January 2009, with initial numbers taken from the assessment and an assumed catch for that year. The initial numbers were made noisy using the same model as for observations elsewhere. F was modified for the first time in 2011, while the decided F in 2010 was equal to the realized F in 2009.

The model was set up to with the following parameter values:
-F reduction: 15, 20 and $25 \%$ relative to F2009
-CVmod; CV of the survey estimate of SSB: 0 and 0.3
-CVi; CV of "the random year multiplier" reflecting the implementation error in a given year: 0 and 0.3
-CVage; CV of the implementation error by age, reflecting the variability in selection at age, was set to 0.2 for all ages in all years. This value is also used for the uncertainty of the initial stock numbers.

## F-based simulations

## Choice of biological input

A time series of estimated recreational catch at age was presented to the 2010 AFWG. These catches show rather low variation (13-16 kt) over the years 1984-2009, while commercial catch has varied between 22 and 75 kt . The recreational catches thus represent a considerable fraction of total catch, particular in recent years (about 35\%) when commercial catch has been low. The new catch data including recreational catches is expected to better reflect the stock dynamics, but the uncertainty of these additional catches are considered to be large. It is, therefore, useful to explore whether the simulation results are sensitive to the differences in the two sets of input data.

In addition, there is large uncertainty relating to the current and future recruitment regime. The long term decline in recruitment, also in periods with high SSB, could indicate that the high recruitments experienced in the early history may not reoccur even with a rebuilt stock. Therefore, two versions of hockey-stick S-R relationships were assumed for each of the time series; one utilizing the full time series, thereby allowing for high R at high SSB, and one based on the most recent years when only low R has been observed.

The input data based on the new catch data, combined with the assumption of low future recruitment was considered to be the most realistic basis for simulations. Si mulations with other inputs were run to examine the robustness of the conclusions against these differences of input.

Equilibrium estimates as function of F for yield, SSB and R is estimated as a part of the HCS simulation. Thereby, also F0.1, Fmax, F35\%SPR and Fcrash are estimated. The results of these 4 break point analysis are shown below. The Fmax was at Fcrash (bey ond Fcrash if estimated by classical Yield per Recruit analyses) in all cases, except the one with added catch and high recruitment, where it was just below Fcrash.

## "Reported catch"

A break point analysis of the full series (red line in Figure 8) gives a plaeau of 33.3 million recruits above an ssb of 114 kt . Assuming changed recruitment conditions since 2000 (Figure 8; the green line based on the lower 6 points) gives a plateau at 17.6 mill recruits above an ssb of 71 kt .


Figure 8.Stock-Recruit plot for the year-classes 1984-2005. Red line fitted to full time series, green line fitted to the 6 lowest points (year-classes 2000-2005). From an assessment based on commercial catch.

Simulations with hockey-stick recruitment functions with these breakpoints were run, setting recruitment variation according to a $\mathrm{cv}=0.33$ for the $\log$ residuals. This give the following results

| R- <br> plateau | bp <br> SSB | Fcrash | F0.1 | Y at <br> $\mathrm{F}_{\text {crash }}$ | Y at <br> $\mathrm{F}_{\max }$ | Y at <br> $\mathrm{F}_{0.1}$ | SSB at <br> $\mathrm{F}_{\text {crash }}$ | SSB at <br> $\mathrm{F}_{\max }$ | SSB at <br> $\mathrm{F}_{0.1}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 17.6 | 71 | 0.30 | 0.17 | 27 | $27^{*}$ | 25 | 72 | $72^{*}$ | 125 |
| 33.3 | 114 | 0.35 | 0.17 | 52 | $52^{*}$ | 48 | 114 | $114^{*}$ | 237 |

The F at $35 \% \mathrm{SPR}=0.16$, which is very close to F 0.1 . This shows slightly higher Fcrash values and slightly lower F0.1 than those calculated at WKPOOR2 (ICES 2009). The changes in Fcrash seem to be mainly caused by the changed maturity input for the historic SSB time series. The change in F0.1 is related to updated values for exploitation pattern and weights at age.

## Recreational fisheries added

The SSB and R time series from an assessment based on the added catch data gives the S-R relationship shown in Figure 9. A break point analysis of the full series (red line) gives a plaeau of 46.7 million recruits above an ssb of 139 kt . Assuming changed recruitment conditions since 2000 (the green line based on the lower 6 points) gives a plateau at 28.6 mill recruits above an ssb of 103 kt .


Figure 9. Stock-Recruit plot for the year classes 1984-2005. Red line fitted to full time series, green line fitted to the 7 lowest points (year classes 2000-2005). From an assessment based where recreational catches are added.

HCS analysis with hockey-stick recruitment functions with these breakpoints were run, setting recruitment variation according to a $\mathrm{CV}=0.33$ for the $\log$ residuals. This give the following results

| R- <br> plateau | bp <br> SSB | Fcrash | F0.1 | Y at <br> $\mathrm{F}_{\text {crash }}$ | Y at <br> $\mathrm{F}_{\text {max }}$ | Y at <br> $\mathrm{F}_{0.1}$ | SSB at <br> $\mathrm{F}_{\text {crash }}$ | SSB at <br> $\mathrm{F}_{\text {max }}$ | SSB at <br> $\mathrm{F}_{0.1}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 28.6 | 103 | 0.32 | 0.16 | 45 | $45^{*}$ | 41 | 104 | $104^{*}$ | 209 |
| 46.7 | 139 | 0.38 | 0.16 | 73 | 74 | 67 | 139 | 158 | 341 |

The F at $35 \% \mathrm{SPR}=0.15$, which is very close to F0.1. The values of F0.1 are similar to those calculated including only "reported" catch. A minor increase for Fcrash seem to be mainly a result of the added catches being larger relative to total catch in recent years, thereby causing a larger relative increase for the recent $R$ and SSB values compared to the older. These new Fcrash values are still close to the current F, and there is still very important to reduce the F well below the current level. Thus, the main conclusions with these new data are similar to those given by WKPOOR2 (ICES 2009).

Precautionary Approach Reference points have not been established for this stock. Taking this "added catch and low break point" data set as the best description of the stock dynamic, the break-point SSB=103 kt would be a good candidate for Blim, and Fcrash $=0.32$ a good candidate for Flim.
Bpa would then be the estimated SSB which has high probability of being above the break-point SSB ( 103 kt ) and Fpa the estimated F that has high probability of being below Fcrash (0.32). The simulations based on this break point (Figures 10 and 11) shows that the 90 percentile of F falls below Fcrash when average F falls below about 0.16 . This reflects, however, the frequency distribution of realized Fs corresponding
to the rebuilding conditions and the assumed errors. This can thus not be taken as a candidate Fpa. The simulation setup applied is not well designed for giving estimates of PA values.

## Simulations based on alternative stock dynamics and regulations

The simulations based on added catch and low recruitment are presented in some detail in the section below. The three other sets of biological input data described above were also tested out and some results are summarized in Tables 4-9. The main patterns are similar to those described in the section below. Those based on commercial catches were generally a bit more pessimistic than those based on added catch. Those with high recruitment were a bit more optimistic than those with low, not only in terms of biomass and catches, but also in terms of rebuilding time (Table 9).
Simulations with the TAC versions of the software are also shown in Tables 8 and 9 . These are indicating slightly slower rebuilding. The main reason is probably that the CVobs used here results in some cases where the stock is strongly overestimated and the corresponding quota results in a realized F much higher than intended. This might illustrate that when uncertainty is high, pure TAC-regulation could be more risky than a F control regime.

## Simulations based on the added catch data and low recruitment

The current F (in 2008 and 2009) was in the assessment estimated to be near 0.30 which is close to the average of the previous 5 yrs. The selection and weights at age was set as the recent 5 yr average in the vpa. In the forward simulations the total catch in 2009 was set equal to 2008 ( 37 kt ), and the recovery plan started from 2010 with an intended F equal to the realized F in 2009. Three reduction rates; $15 \%, 20 \%$ and $25 \%$ was simulated. The rebuilding target (a survey value of 60 kt spawners $=95-98$ average) was in the model replaced by the $95-98$ average of ssb in the historic assessment $=180 \mathrm{kt}$. This corresponds to scaling the survey to the historic vpa, while keeping the assumed CV. Various assumptions were explored. The CV of the implementation error (CVi) and the CV of the survey (CVmod) were both run for the values 0 and 0.3 . The auto correlation for the observation model (survey) was kept at 0 . The survey is recognized as uncertain, and it is rather obvious that it will be very difficult to design regulations that will give the desired F-reduction by a high accuracy. CVs at 0.3 could be considered realistic. Some results with these CVs set to zero are shown for comparison. An additional "CV at age" (uncertainty in age distribution) was set to 0.2 for all ages in all simulations, also those where other CVs were zero.
No attempts have been made to split the simulation on the commercial and the recreational fleet. If the additional regulations in coming years work equally on the two fleets, it should be expected that about $35 \%$ of catches will be taken in the recreational fishery.
Figures 10-13 show the time development for the average of 1000 simulations. Figures 14 and 15 show the development of 20 individual trajectories for F and SSB. This illustrates the larger variability caused higher CVs. Tables 4-6 summarize some of the key findings, and compare these simulations with those based on other input and assumptions. In the further description the terms $0.15,0.20$ and 0.25 are used for 15 , 20 and $25 \%$ F-reduction steps. This is done to avoid confusion with percentiles in the observed frequency distributions.

The average F in 2010 comes out higher when errors in observation and implementation are assumed (Figure 10). The further propagation of average F is rather similar
with and without those errors. With errors assumed the average F levels off at 0.11 , just above the minimum threshold. The 90 percentile of F is below the Fcrash already in 2013 when CVs are zero, while for the 0.15 curve with CVs of 0.3 this does not occur before 2017 (Figure 11).
For all F-reduction rates the average catches decrease to a minimum of 23 kt , and increases from there (Figure 12). Both the decrease and the increase are slowest for the 0.15 case.

Average SSB increases very slowly the first 4 years (Figure 13a). The 0.15 curve shows an average SSB hitting the rebuilding target ( 180 kt ) in 2025. The 10 percentile of the SSB distribution (Figure 13b) does not reach the rebuilding target within the period considered, but it crosses Blim in 2026 in the 0.15 case, while for the 0.25 case it occurs 3 years earlier. The lower panel of Figure 13b shows the probability of SSB<Blim. In the 0.15 case the curve falls below the $10 \%$-line in 2026 as also illustrated by the upper panel in Figure 13b.


Figure 10. Average realized $F$ when the reduction steps in "action years" are $\mathbf{1 5 \%}$ (blue), $\mathbf{2 0 \%}$ (red) and 25\% (green)



Figure 11. The 90 percentile for the distribution of realized Fs when the reduction steps in "action years" are $\mathbf{1 5 \%}$ (blue), $20 \%$ (red) and $25 \%$ (green). Dotted red line is the Fcrash.


Figure 12. Average realized catch when the reduction steps in "action years" are $\mathbf{1 5 \%}$ (blue), 20\% (red) and $25 \%$ (green).


Figure 13a. Average SSB when the reduction steps in "action years" are $\mathbf{1 5 \%}$ (blue), $\mathbf{2 0 \%}$ (red) and 25\% (green).


Figure 13b. Upper panel; The 10 percentile for the distribution of SSBs when the reduction steps in "action years" are $15 \%$ (blue), $20 \%$ (red) and $25 \%$ (green). Dotted red line is the Blim. Lower panel: The probability that $\mathrm{SSB}<\mathrm{Blim}$, when the reduction steps in "action years" are $15 \%$ (blue), $20 \%$ (red) and $25 \%$ (green).



Figure 14. Trajectories of F for 20 individual simulations with $15 \% \mathrm{~F}$ steps in "action years".


Figure 15. Trajectories of SSB for 20 individual simulations with $15 \%$ F steps in "action years"

## Summary of the simulations

For the proposed 0.15 version of the plan the average SSB reaches the rebuilding target in 2024 , which corresponds to about $50 \%$ probability for the true stock to be above in that year. High probability towards Blim is obtained 2 years later. True F safely below F crash (Flim) is obtained in 2017. This is an important milestone in the sense that the management then has reached a safe ground, and further reductions of F will gradually increase the SSB until equilibrium is reached for Fs close to 0.1. The equilibrium SSB for $\mathrm{F}=0.11$ is (without density dependant effects) estimated at 280 kt .
A faster route to safe ground is obtained by having larger F steps. F safely below Fcrash occurs in 2016 in the 0.2 version, and in 2015 in the 0.25 version (Table 9). Another approach could be to have the F reduction unconditional to the survey, at least for the first few years. Simulations applying reductions every year until the lower limit is reached are summarized in Table 7 ( $3^{\text {rd }}$ column). Then F safely below Fcrash occurs in 2014 in the 0.15 version, and in 2013 in the 0.2 and 0.25 versions.

## Conclusion

If the plan is fully implemented it will lead to a safe rebuilding. Under presumed realistic errors a rather long rebuilding period is required, but the fishing mortality comes down to fairly safe levels within few years. On this basis the proposed rule is considered to be in accordance with the Precationary Approch. Increasing the F step, or aiming for annual reductions unconditional to survey results during the first 3-5 years, will both contribute to a faster and safer rebuilding. If future observations show recruitment declines stronger than assumed in the current stock-recruit model, the plan may need revisions. The new data on recreational fisheries also highlights the need to consider further regulations on these activities to obtain the F-reductions specified in the plan.

The current regulations aiming for protection of local stock components should be maintained. This should beimproved when the scientific basis is improved.

## Other relevant issues raised by the request

In these quantitative simulations and analyses no direct attempts have been made to take account of the stock complexity. Genetic studies indicate that the cod in some fjords could be separate stocks isolated from neighboring stocks. An assessment of the merged stock is not likely to detect fluctuations of the smaller components, and thereby the current assessment approach involves some risk to local stocks. The stock complex is still not fully mapped, but the existence of local stocks also calls for special attention for protecting genetic diversity. Full monitoring and research on small local stocks requires large efforts and may not be realistic. A possible approach could be to obtain information from local fisheries and look for data that could be appropriate indicators for at least detecting sharp declines of local stocks. The established strategy of more strict regulations inside the fords than outside should be continued.

A fixed natural mortality of 0.2 is used both in the assessment and the simulations. Some fjord studies (Pedersen and Pope, 2003a and b, Mortensen 2007, Pedersen et al., 2007, Aas, 2007). indicate that the main predators on young cod is larger cod, cormorants and saithe. There are no estimates of annual predation mortality for the stock complex. Thus, the development of the cod predators, mentioned in the request, is not taken into account. Reduced predator stocks may enhance the rebuilding of cod, while an increase of predators may inhibit the process and require prolonged strong regulations of the fishery for obtaining the rebuilding target.

Table 4. F-based simulations. $15 \%$ reduction, conditional to survey results. The shaded part specifies the input and reference values. The unshaded part refers to the year when the criterion specified in the left column is reached. The bold heading indicates the input considered to be most realistic. The bold rows relate to PA -criteria.

| $15 \%$ F-red <br> CVi=CVobs $=0.3$ | Commercial catch, low bp | Commercial catch, high bp | Added catch, low bp | Added catch, high bp |
| :---: | :---: | :---: | :---: | :---: |
| Breakp. SSB | 71 | 114 | 103 | 139 |
| TargetSSB | 139 | 139 | 180 | 180 |
| R-plateau | 17.6 | 33.3 | 29.1 | 46.7 |
| Fcrash | 0.30 | 0.35 | 0.32 | 0.38 |
| F0.1 | 0.17 | 0.17 | 0.16 | 0.16 |
| avF<Fcrash | 2015 | 2014 | 2012 | 2011 |
| $\mathrm{P}(\mathrm{F}>$ Fcrash $)<0.1$ | 2019 | 2018 | 2017 | 2016 |
| avF<F0.1 | 2019 | 2019 | 2019 | 2019 |
| avSSB>bp | 2021 | 2024 | 2015 | 2017 |
| avSSB>Btarget | >2029 | 2029 | 2020 | 2020 |
| $\mathrm{P}(\mathrm{SSB}<\mathrm{bp})<0.1$ | >2029 | >2029 | 2026 | 2026 |

Table 5. F-based simulations. $20 \%$ reduction, conditional to survey results. The shaded part specifies the input and reference values. The unshaded part refers to the year when the criteria specified in the left column is reached. The bold heading indicates the input considered to be most realistic. The bold rows relate toPA-criteria.

| 20\% F-red <br> CVi=CVobs=0.3 | Commercial catch, low bp | Commercial catch, high bp | Added catch, low bp | Added catch, high bp |
| :---: | :---: | :---: | :---: | :---: |
| Breakp. SSB | 71 | 114 | 103 | 139 |
| TargetSSB | 139 | 139 | 180 | 180 |
| R-plateau | 17.6 | 33.3 | 29.1 | 46.7 |
| Fcrash | 0.30 | 0.35 | 0.32 | 0.38 |
| F0.1 | 0.17 | 0.17 | 0.16 | 0.16 |
| avF<Fcrash | 2014 | 2013 | 2012 | 2016 |
| $\mathrm{P}(\mathrm{F}>$ Fcrash $)<0.1$ | 2017 | 2016 | 2016 | 2015 |
| avF<F0.1 | 2016 | 2017 | 2017 | 2017 |
| avSSB>bp | 2019 | 2022 | 2014 | 2016 |
| avSSB>Btarget | >2029 | 2024 | 2022 | 2019 |
| $\mathrm{P}(\mathrm{SSB}<\mathrm{bp})<0.1$ | >2029 | >2029 | 2024 | 2024 |

Table 6. F-based simulations. $25 \%$ reduction, conditional to survey results. The shaded part specifies the input and reference values. The unshaded part refers to the year when the criterion specified in the left column is reached. The bold heading indicates the input considered to be most realistic. The bold rows relate toPA-criteria.

| 25\% F-red <br> CVi=CVobs=0.3 | Commercial <br> catch, low bp | Commercial <br> catch, high bp | Added catch, <br> low bp | Added catch, <br> high bp |
| :--- | :---: | :---: | :---: | :---: |
| Breakp. SSB | 71 | 114 | 103 | 139 |
| TargetSSB | 139 | 139 | 180 | 180 |
| R-plateau | 17.6 | 33.3 | 29.1 | 46.7 |
| Fcrash | 0.30 | 0.35 | 0.32 | 0.38 |
| F0.1 | 0.17 | 0.17 | 0.16 | 0.16 |
| avF<Fcrash | 2016 | 2013 | 2012 | 2012 |
| P(F>Fcrash)<0.1 | $\mathbf{2 0 2 0}$ | $\mathbf{2 0 1 6}$ | $\mathbf{2 0 1 5}$ | $\mathbf{2 0 1 4}$ |
| avF<F0.1 | 2016 | 2016 | 2016 | 2016 |
| avSSB>bp | 2018 | 2021 | 2014 | 2016 |
| avSSB>Btarget | $>2029$ | 2026 | 2021 | 2018 |
| P(SSB<bp)<0.1 | $\mathbf{2 0 2 9}$ | $\mathbf{2 0 2 9}$ | $\mathbf{2 0 2 3}$ | $\mathbf{2 0 2 3}$ |

Table 7. F-based simulations. 15\% reduction, unconditional to survey results. The shaded part specifies the input and reference values. The unshaded part refers to the year when the criterion specified in the left column is reached. The bold heading indicates the input considered to be most realistic. The bold rows relate to PA-criteria.

| 15\% F-red <br> CVi=CVobs=0.3 | Commercial <br> catch, low bp | Commercial <br> catch, high bp | Added catch, <br> low bp | Added catch, <br> high bp |
| :--- | :---: | :---: | :---: | :---: |
| Breakp. SSB | 71 | 114 | 103 | 139 |
| TargetSSB | 139 | 139 | 180 | 180 |
| R-plateau | 17.6 | 33.3 | 29.1 | 46.7 |
| Fcrash | 0.30 | 0.35 | 0.32 | 0.38 |
| F0.1 | 0.17 | 2013 | 2013 | 2016 |
| avF<Fcrash | 2015 | 2014 | 2014 | 0.16 |
| P(F>Fcrash)<0.1 | 2015 | 2016 | 2014 | 2011 |
| avF<F0.1 | 2017 | 2026 | 2020 | 2014 |
| avSSB>bp | 2029 | 2024 | 2021 | 2015 |
| avSSB>Btarget |  |  | 2022 |  |
| P(SSB<bp)<0.1 | 2027 |  | 2 |  |

Table 8. TAC-based simulations (Intended F translated to TAC, realized catch lead to the realized F). $15 \%$ reduction, conditional to survey results. The shaded part specifies the input and reference values. The unshaded part refers to the year when the criterion specified in the left column is reached. The bold heading indicates the input considered to be most realistic. The bold rows relate toPA-criteria.

| 15\% F-red <br> CVi=CVobs=0.3 | Commercial <br> catch, low bp | Commercial <br> catch, high bp | Added catch, <br> low bp | Added catch, <br> high bp |
| :--- | :---: | :---: | :---: | :---: |
| Breakp. SSB | 71 | 114 | 103 | 139 |
| TargetSSB | 139 | 139 | 180 | 180 |
| R-plateau | 17.6 | 33.3 | 29.1 | 46.7 |
| Fcrash | 0.30 | 0.35 | 0.32 | 0.38 |
| F0.1 | 0.17 | 0.17 | 0.16 | 0.16 |
| avF<Fcrash | 2018 | 2017 | 2015 | 2014 |
| P(F>Fcrash) $<\mathbf{0 . 1}$ | $\mathbf{2 0 2 2}$ | $\mathbf{2 0 2 2}$ | $\mathbf{2 0 2 1}$ | $\mathbf{2 0 2 0}$ |
| avF<F0.1 | 2022 | 2023 | 2022 | 2023 |
| avSSB $>$ Btarget | 2024 | 2029 | 2015 | 2018 |
| SSB $>$ Btrig | $\mathbf{> 2 0 2 9}$ | $\mathbf{> 2 0 2 9}$ | $\mathbf{2 0 2 8}$ | $\mathbf{2 0 2 2}$ |
| P(SSB<bp)<0.1 | $\mathbf{2 0 2 8}$ | $\mathbf{> 2 0 2 9}$ | $\mathbf{> 2 0 2 9}$ | $\mathbf{> 2 0 2 9}$ |

*Hardly obtainable: Btarget close to equlibr SSB at $\mathrm{F}=0.1$

Table 9. Overview of all simulation runs. Combinations of model versions, assumptions, input data and CVs. "Low" break point is based on year-classes 2000-2005. "High" is based on full time series. The 4 columns to the right show the year ( 20 xx ) when P ( F above Fcrash)<0.10.

|  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

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## Annex 11 - Technical Minutes of a review of the ICES Arctic Fisheries Working Group (AFWG)

Report 2010 (by correspondence)
6-20 May 2010

| Reviewers: | Noel Cadigan (chair) |
| :--- | :--- |
|  | Rasmus Nielsen |
|  | Jean-Claude Mahe |
| Chair WG: | Fernando González |
| Secretariat: | Bjarte Bogstad |
|  | Mette Bertelsen |

Audience to write for: advice drafting group, ACOM, benchmark groups and next years $E G$.

## General

Use bullet points and subheadings (Recommendations, General remarks for WGs, etc.) if needed

The Review Group ToR's (from Guidelines for Review Groups):

1. Thoroughly check to ensure that the assessment is carried out according to the descriptions in the stock annex.
2. Check the content of figures and tables, and review whether the texts are supported by the scientific results.
3. Check consistency with the previous years reporting.

Reviewers also provided comments additional to the ToR's, and these are included as well.

The Review Group considered the following stocks:

- Cod in Subareas I and II (Northeast Arctic cod)
- Cod in Subareas I and II (Norwegian coastal cod)
- Greenland halibut in Subareas I and II
- Haddock in Subareas I and II (Northeast Arctic)
- Saithe in Subareas I and II (Northeast Arctic)
- Beaked Redfish (Sebastes mentella) in Subareas I and II
- Golden Redfish (Sebastes marinus) in Subareas I and II
- Capelin in Subareas I and II (Barents Sea), excluding Division Ila west of $5^{\circ} \mathrm{W}$

And the following special requests:

- none

The RG acknowledges the intense effort expended by the working group to produce the report. The report was well structured and information was usually easy to find.

The stocks listed above were all updates, except Greenland halibut in Subareas I and II (SALY), and were reviewed by the group. In most cases (except Barents Sea capelin) a quality handbook $w$ as available with instructions on the procedure to carry out the assessment.

The reviewers met by correspondence and had contact through e-mail and sharepoint. For the purpose of evaluation the chair of the review group split the stocks between the review ers. The chair read all stock reports. It was checked by the review ers whether the procedures followed were according the procedures established in a previous bench mark assessment. In most cases the present assessments were also compared with those of last year. There was insufficient time to review other chapters of the report. Also no draft stock summaries were considered by the review group.

| Stock | Name | Assessment <br> Type | Reviewers (1st) |
| :--- | :--- | :--- | :--- | :--- |
| cod-arct | Cod in Subareas I and II (Northeast Arctic) | Update | Rasmus Nielsen |
| cod-coas | Cod in Subareas I and II (Norwe gian <br> coastal waters) | Update | Rasmus Nielsen |
| had-arct | Haddock in Subareas I and II (Northeast <br> Arctic) | Update | Rasmus Nielsen |
| sai-arct | Saithe in Subareas I and II (Northeast <br> Arctic | Update | Rasmus Nielsen |
| cap-bars | Capelin in Subareas Iand II (Barents Sea), <br> excluding Division IIa west of 5${ }^{\circ}$ W | Update | Noel Cadigan |
| ghl-arct | Greenland halibut in Sub-a reas I \& II Sa me advice as <br> last year <br> smn-arct Red fish Sebastes mentella Subareas Iand IIUpdate | Jean-Claude Mahe |  |
| smr-arct | Red fish Sebastes marinus Subareas Iand II | Same advice as <br> last year | Jean-Cla ude Mahe |

## Stock: Cod in subareas I and II (Norwegian coastal waters) (report section 2)

Short description of the assessment: extremely useful for reference of ACOM!

1) Assessment type: Schedule says update, but the assessment was in many ways new.
2) Assessment: Analytical.
3) Forecast: Type: MFDP Program.
a. Catch options for 2011 and 2012. Reducing F4-7 by 15\% each year over two years will result in an increase in SSB.
b. With $15 \%$ reduction steps for F , it will take 7 years to have high probability that F < Fcrash (see below), 10 years for having average SSB $>$ the stock rebuilding plan target, and about 15 years to have high probability for SSB>Blim.
4) Assessment model:
a. XSA- including commercial and recreational catches and tuning by one acoustic survey, and standard VPA (SVPA).
b. Retrospective analyses showed a tendency for XSA to overestimate F4-7 in the last year. XSA F's were combined with F's based on an external analysis presented in an annex to generate F's to use in the final SVPA. This procedure was not described clearly.
5) Consistency:
a. The assessment gives the same perception of $\mathrm{SSB}, \mathrm{R}$ and F as last years assessment.
b. In the two last years a SURBA assessment was used while an XSA assessment is used this year. In earlier years the w orking group has not considered the XSA to give reliable results. This year the working group did not consider SURBA to give reliable results.
6) Stock status:
a. The assessment proposed candidates for $\mathrm{Blim}_{\lim }$ and $\mathrm{Flim}: \mathrm{Blim}_{\mathrm{lim}}=103 \mathrm{Kt}$ SSB; Flim $=0.32$ (Fcrash). B2009 ( 80 Kt ) < Blim; F2009 (0.31) < Flim.
b. F decreased in 1999-2000 and has since been relatively stable. F2009 = 0.31 .
c. F0.1 is estimated to be 0.16 , and a safe long term Fmsy-target is, according to the w orking group, close to 0.16. The corresponding MSY Btrigger will be in the range of 150-200 kt.
d. SSB is estimated to be stable but at low level, and well below long term average.
e. Survey estimates of age group 1-3 in 2009 are among the lowest in time series and indicate low recruitment.
7) Man. Plan.:
a. There is no management plan.
b. Fishery is managed with annual TAC and technical measures. Since 2001 the TAC has been 21 Kt .
c. For 2010, an additional TAC of 10 kt was set for the recreational and tourist fishery and added to the commercial fishery TAC of 21 kt .
d. Since 2004, the ICES advice has been no catch and a recovery plan should be implemented.
e. A rebuilding plan has been proposed by Norwegian authorities and has been evaluated by the working group.

## General comments

This is a well ordered section. Some parts of the text in the report are updated from last year's assessment based on updated analyses; however, major revisions are presented. Last year a SURBA assessment was used. This year an XSA assessment is used with additional inclusion of new catch at age time series as well as introduction of fixed maturity ogive in the assessment. The working group reasoning for the revisions have been to have some additional basis for evaluating the proposed rebuilding plan for the stock. Last years assessment was considered tentative and used as a basis for advice last year.

A new catch series for recreational and tourist fisheries is added to the commercial catch. This time series contributes around $1 / 3$ of the catch. Major assumptions are made in relation to estimating catch in this new time series (see below). The combined data series is by the working group found to fit better with the survey, using stock dependent catchability for ages 2 and 3 in a XSA. The XSA is otherwise set to standard values given in the Quality Handbook Stock Annex (also from last year). General information regarding the stock and earlier assessments are given in an updated Quality Handbook Stock Annex. As such the update cannot be considered as an actual SPALY.

There was little discussion of the quality of the catch at age data. Sources for large errors include unreported catch (i.e. recreational+tourism) and catches mixed with the NEA stock. An accurate resolution to these problems may not be possible; however, it would be useful for the w orking to assess, perhaps using only expert opinion, the accuracy of the catch. If the plausible catch times series have substantial pointwise range then this uncertainty should be accounted for in the assessment.

Furthermore, from the retrospective plots of the XSA it is stated that there is a tendency for the F4-7 in the last year to be overestimated, and this is dealt with in the final SVPA by setting a terminal F based on external analysis presented in an Annex. Also, this is a change compared to last year.

The maturity in the surveys are variable, but show overall a declining tendency. Maturity data originates from surveys. The ogives are variable which according to the working group is influenced by time of surveying more than real between year variation. On this basis, the working group conclude that it might be better to use a fixed (average) maturity ogive over years. This variability has not been analysed in detail. A fixed maturity ogive has been implemented in this years final XSA assessment compared to previous years (survey based) assessment using SURBA. Further benchmark investigations of this is needed.

Results and full comparative plots of exploratory runs testing each of these changes in the assessment individually has not been performed, i.e. with inclusion/exclusion of the new data time series, different maturity ogives, and different assessment mod-
els. On this basis, the review er found it difficult to evaluate the impact of each of the individual changes made in the assessment, as well as difficult to evaluate the impact of changing the assessment model. Thorough benchmark analysis should have been (and needs to be) done to evaluate the changes in the assessment.

Several exploratory runs with different methods and input data were carried out this year, and the most "appropriate" was selected by the working group. The basis for appropriateness has not been benchmark analysed. The assessment should have presented comparative plots for SSB, Fbar and Recruitment for each of the assessment runs compared to last years assessment. This would have eased the evaluation of the impact of each of the assessments with variable input data, settings and model use.

The outcome of the tentative assessments gives the same perception of the development of stock and fishery as last year's assessments.

## Technical comments

1 ) The review was restricted to a check whether the procedures described in the technical annex (handbook) were applied, and a comparison with last years assessment was performed. The technical annex (quality handbook) has been revised this year and the procedures of the new assessment this year has been included. The present assessment must be characterized as a new assessment which needs to follow standard benchmarking procedures.
2 ) Also a comparison with last year's report was made. This is commented in detail above and below.

3 ) No checks on the calculation of the international age structure have been carried out by the reviewer.
4 ) The assessment with SURBA is based on an acoustic survey. Because cod contributes only a relatively low fraction of the observed acoustic values, the estimates of the survey are more sensitive to allocation error. The WG is aware of this. This contributes to uncertainty in the point estimates of the analyses but not to the perception of the present stock size.
5 ) It would be an advantage to show survey trends in CPUE normalized to unit mean in Figures 2.7-2.13.

6 ) There should in section 2.3.1 be made reference to both Table 2.1.13 and 2.1.14.

7 ) There is no indication in the text or table for which time of the SSB is calculated. Biomass and landings input in the prediction is similar to 2009 biomass and landings.
8 ) It would have been a benefit for the review if the assessment had presented standard stock overview plots with Landings, Fishing Mortality, Recruitment and SSB, as well as a standard Stock-Recruitment plot, a Y/R plot and a precautionary approach plot (which traditionally are associated to the advisory sheets).
9 ) The methods used to infer recreational and tourist catch are based on little actual data and must be considered tenuous. The WG estimates of these components comprise approximately $30 \%$ of the total catch which is a substantial amount. Error in the catch at age should be accounted for in the assessment model.

10 )Age distribution from the commercial fishery long line and hand line fishery is applied to these recreational fishery catch time series. This assumption has not been evaluated.

11 )All recreational fishery cod catches are assumed to be coastal cod (i.e. no fraction of NEA cod).
12 )Last year an old version of SURBA (version 2.1) was used. The catchabilities calculated by this version of SURBA are not correct.
13 )XSA's are not converged. This was considered in the 2009 report of WGMG.

## Conclusions

The assessment has been performed correctly. There is need for a benchmark in the short time. The information given by the assessments is sufficient to provide advice.

The final stock assessment gives the same perception of SSB, R and F development as last years assessment.

## Stock: North-East Arctic Cod (Subareas I and II) (report section 3)

Short description of the assessment: extremely useful for reference of ACOM!

1) Assessment type: update/SPALY
2) Assessment: Analytical.
3) Forecast: Type: MFDP Program. Fbar (5-10) in 2010 equal to same Fbar in 2009. Exploitation pattern: Average of 2007-2009.
4) Assessment model:
a. XSA using 4 tuning fleets ( 3 surveys and 1 commercial cpue) as an up-date assessment with identical settings to last years assessment;
b. maturity data and weight at age in the stock data are from surveys;
c. M is set as the value M 1 of 0.2 (by age and year) but added a matrix of natural mortality caused by cannibalism (M2);
d. Unreported catches are considered low by the WG in the most recent years, and in 2009 IUU catches was not included.
e. Additional models presented were TISVPA, Gadget, and survey calibrated VPA.
f. There is general consistency in trends and perception of the stock dynamics between the models.
5) Consistency: Last years assessment was accepted and used as a basis for advice. This year's assessment is consistent with the assessment for the two last years.
6) Stock status:
a. The stock is within safe biological limits. Fsq<Fpa (and well below Flim, lowest $F(5-10)$ since 1990) and SSB $>$ Bpa (and well above Blim, SSB at 1145 kt is the highest since 1947).
b. Also, recruitment in recent years ( 589 millions in 2009) is around long term average ( 610 millions).
c. Reference points have not been revised since 2003. Target reference points according to MSY has been evaluated including cannibalism and a segmented regression SSB-recruitment relationship. The results indicated that a long term yield is fairly stable for a range of fishing mortalities between 0.25 and 0.6 . Density dependent effects in cannibalism and growth are by the WG considered as the main reason for this rather wide F-range with stable high yield
7) Man. Plan.
a. There is an agreed management plan. An amended HCR from 2009 was evaluated and was considered by the WG to be in accordance with the precautionary approach.
b. The fishery is regulated through TAC quota with max. $10 \%$ change. Other technical measures are described in the stock annex.
c. The 2009 TAC was not taken in 2009 , as catches were slightly below .
d. The fishery is mixed with Norwegian coastal cod fishery, and there are by-catches of NEA cod in the shrimp fishery.

## General comments

This was a well documented, well ordered and well considered section. The text in the report is an update from last year's report. The various analytic assessments give the perception of a significant increasing stock as a result of a reduction in fishing mortality.

## Technical comments

1 ) The review was restricted to a check of whether the procedures described in the technical annex (handbook) were applied. This was the case. No deviations were spotted. Little or no attention has been given to the additional models by the reviewer. These additional models are not described in the annex.

2 ) Also, a comparison with the assessment in last year's report was made. The procedures used were the same as last year. The results of the assessment are in line with last year's assessment.
3 ) The results of the XSA assessment were robust to assumptions made on $q$ (stock size dependent catchability) for age groups 3-7 as well as on sensitivity to the length of the tuning period. This was tested by the WG. The final XSA run was compared with single tuning fleet runs with standard shrinkage settings. The swept area bottom trawl estimates from the joint Norwegian-Russian ecosystem survey have been tested as a new survey tuning time series in these single fleet runs.
4 ) There was no description of the methods used to account for cannibalism in the Annex. There was insufficient details presented in the report, and insufficient time available, to review this aspect to evaluate technical correctness.

5 ) Some of the survey indices have been multiplied by a factor 10 . According to the working group, this has been done in order to keep survey dynamics even for very low indices because XSA adds 1.0 to the indices before the logarithm is taken.
6 ) There seems to be a tendency to a general higher level in the log catchability residuals for later years in the Russian bottom trawl survey, as well as a positive trend in the later years for the winter bottom trawl survey (Fig. 3.2).

7 ) There seems to be a small tendency to overestimate F and underestimate SSB when all fleets are included compared to single fleet runs which probably is due to higher influence of shrinkage in single fleet runs. Because it is low, it should not be considered to be problematic. Also, a test run without stock dependent catchability for age groups 3-5 gave slightly lower F and higher SSB for 2009.
8 ) There seems to be goods consistency between the different surveys. Survivors estimates from single fleet runs for all ages are in relatively good agreement between fleets. However, it would be an advantage to have a concluding paragraph on the degree of consistency between fleets and temporal development herein under sections 3.2.1 and 3.2.2.

9 ) Although the XSA is the standard method accepted in the benchmark, the TISVPA is run as an alternative. The same settings as last year was used in
the TISVPA. The two models give same perception of development in the stock, and the TISVPA output is consistent with last years results using the method. As such the results of XSA and TISVPA are very similar. However, in the most recent years TISVPA gives higher SB and SSB compared to XSA, and a slightly lower estimate of $\mathrm{F}(5-10)$. This was also the case for the two last years assessments. Last year, a different loss function was chosen to find a minimum compared to year before. The reviewer evaluated last year that such an approach would be difficult to accept in a benchmark procedure since this may lead to great differences in the results of the assessment between years. Also, the reviewer noted that i) the TISVPA is useful to demonstrate inconsistencies in the catch at age matrix, and ii) the reason for considering a TISVPA would be: suspecting an effect of cohort(size) on the exploitation.
10 )The text in section 3.1 mentions a total landing in relation to Table 3.1a, while the table text to Table3.1a mentions total catch. Table texts for 3.2-3.3 also indicates catch rather than landings of 523431 t , while the text sections writes landings. It is important to be precise here with respect to what is catch and landings, especially when discard is mentioned in the text and partly considered in the assessment.
11 )Sequence in Table numbering is not consistent with the sequence of reference in the text. For example, in relation to Tables 3.4-3.11.
12 )Weight at age in catch is variable between years, periods and countries. Plots of this should be shown in order to be able to follow trends and tendencies (besides Table 3.4). For example, it seems that Norwegian landings weight at age has increased by about 1 kg in the latest period for age groups 6-10, but not in other nation's landings. This should be looked at more closely.
13 )Maturity at age in catch is variable between years, periods and countries. There should be shown plots of this to be able to follow trends and tendencies (besides Table 3.5). There seems to be a quite drastic decrease in the percentage mature at age in 2010 for the Russian data. This should be looked at more closely.
14 )There is no reference to Table 3.8 in the text describing cannibalism (section 3.3.5) even though cannibalism is included in the XSA.
15 )The last part of Table 3.9 is missing - only catch at age numbers up to year 1969 is given here. The full table needs to be included.
16 )With respect to stock weights at age, the text (section 3.3.2) refers to Table 3.12, while stock weights at age is given in Table 3.11.

17 )There are different estimates of unreported catches by Norway and Russia for 2008 in last years assessment. Norway estimated 15 kt in 2008. Russia comes with underutilization of 425 tonnes. In 2008, there were 2 different assessments and prognoses, based on different assumptions on unreported landings. In the 2009 assessment, the (higher) Norwegian estimate was accepted - both for the assessment and prediction. The Norwegian estimates were also used in the past for the final advice. In the 2010 assessment the working group decided not to include unreported catches (IUU). The unreported catches have declined considerable in recent years from over 100 kt in the early 2000's.

18 )This years assessment present considerations of inclusion of data time series for discards in a coming benchmark assessment to meet the reviewer comments from last year.
19 )The catch forecast covers all catches. This means that if any over-fishing takes place the forecasted TAC should be reduced.
20 )The assessment noted a pattern in recent years to have concentrations of cod near the borders of the strata system. This could indicate an increasing amount of fish being distributed outside the strata system. This would indicate a decrease in survey catchability.
21 )An "index ratio by age" method was used to adjust for incomplete Joint Barents Sea winter survey coverage in some years (e.g. 2007). This should be revisited in time, because data before and after years with incomplete coverage can and should be used to fill in the missing data. It would also be desirable to reflect in survey standard errors the additional uncertainty caused by incomplete coverage. This may be more important in the future if more statistically rigorous state-space approaches are used, where process and measurement error are separated and it helps for this to have good information on within-survey error.
22 )Different vessels have been used in the Joint Barents Sea winter survey, but the report provided no information on whether the survey index has been standardized to account for these vessel changes.
23 )XSA is not converged (i.e. Table 3.15). This was considered in the 2009 report of WGMG.
24 )This years assessment includes a discussion of a data time series for discards and a new tuning time series from the swept area bottom trawl joint Norw egian-Russian ecosystem survey, to be considered in an up-coming benchmark assessment.

25 )A benchmark assessment should also consider the apparent increased catchability in the surveys and contradiction in trends of WAAC in landings by different nations.
26 )Last year the reviewer noted: Inspection of historical material indicate a different inter pretation of age of 1st maturity by contemporary age readers. The WG notes (2009) this may affect the SR relationship and biological reference points. This point should also get attention in the next benchmark assessment.

## Conclusions

The assessment has been performed correctly. The information given by the XSA assessment is sufficient to provide advice. The present management has lead to a reduction in F, a substantial increase in SSB and a reduction in unreported landing. There is not an urgent need for a benchmark assessment.

## Stock: North Northeast Arctic Haddock (Subareas I and II) (report section 4)

Short description of the assessment: extremely useful for reference of ACOM!

1) Assessment type: update/SPALY
2) Assessment: Analytical.
3) Forecast: Type: MFDP Program. Analytical forecast is presented.
a. Recruitment is estimated by RCT3 based on age $0-3$ from surveys using same procedure as in previous years.
b. F2010=F2009 because F shows a decreasing tendency from 2007-2009; fishing pattern for 2010-2012 $\mathrm{Fsq}=\mathrm{F}(2007-2009)$ scaled.
c. The $25 \%$ limitation HCR in restricting TAC results in a TAC of 303 kt for 2011 ( $+25 \%$ compared to 2010 TAC of $243 \mathrm{kt)}$ corresponding to $\mathrm{F}=0.31$.
d. The assessment has a retrospective pattern (see below), and one can anticipate a similar pattern in the advice such that in 2-4 years the assessment will say that F in 2010 was somewhat above Fpa, and that the forecasted changes in SSB were not achieved
e. Weight at age and proportion mature is dependent of year class strength, and the WG decided to use smoothed averages for weight at age in stock and catch, maturity, and natural mortality as has been observed in previous years following good recruitment (see section 4.4.4 in WG report).
4) Assessment model:
a. XSA (FLR) using 3 tuning fleets (3 surveys);
b. maturity data are from surveys;
c. M is estimated including predation by cod $(0.2+$ predation mortality by year and age);
d. Unreported catches are considered low by the WG in the most recent years, and in 2009 IUU catches was not included
e. The settings are the same as used in 2009.
5) Consistency: Last years assessment was accepted and used as a basis for advice. This year's assessment is consistent with the assessment for the two last years. There is a tendency to overestimate F and under-estimate SSB in the terminal assessment year.
6) Stock status: The stock is within safe biological limits.
a. Fsq just below $\mathrm{F}_{\mathrm{pa}}$ (and well below Flim ) and $\mathrm{SSB}>\mathrm{B}_{\mathrm{pa}}$ (and well above Blim); SSB is the highest since 1950.
b. Recruitment in recent years is well above long term average - especially the 2006 year class is strong. Surveys indicate a relatively strong 2009 year class as well.
c. Reference points have not been revised since 2000 ( $\mathrm{Blim}_{\mathrm{lim}}=50 \mathrm{kt}, \mathrm{B}_{\mathrm{pa}}=80$ $\mathrm{kt}, \mathrm{Flim}=0.49$ and $\mathrm{F}_{\mathrm{pa}}=0.35$. No work was done this year to revise the reference points for the stock due to time constraints by the WG
7) Man. Plan. There is an agreed management plan.
a. The fishery is regulated through a TAC with max $25 \%$ change and based on a 1-year HCR. Other technical measures are described in the stock annex.
b. In 2007 ICES evaluated the management plan (that was agreed in 2004), and it was found in accordance with the precautionary approach.
c. The HCR is set to a level corresponding to Fpa and a TAC with max $25 \%$ change, and the HCR regulates F if SSB falls below $\mathrm{B}_{\mathrm{pa}}$.
d. The TAC in 2009 of 194 kt was taken with a minor overshoot (landings 200 kt ). TAC in 2010 is 243 kt .
e. Haddock is fished in mixed fisheries and is mainly taken as by-catch in trawl fishery targeting cod. Also, directed haddock fishery with longline and trawl is conducted, and there are set national quotas divided by gear type.

## General comments

The report is well structured and the text is an update from last years report with relative few changes. The assessment show an increasing stock as a result of a reduction in fishing mortality and good recruitment in the last years (especially 2006).

## Technical comments

1 ) The review was restricted to checking whether the procedures described in the technical annex (Stock Annex) were applied. This was the case. No deviations were spotted.
2 ) Also, a comparison with the assessment in last year's report was made. The procedures used were the same as last year. The results of the assessment are in line with last year's assessment
3 ) The RCT3 procedure for predicting recruitment in the short term forecasts is not described well in the annex.
4 ) Catch at ages 1-2 are not used in the assessment, although the annex indicates they should be. Also, survey indices at ages 1-2 are not used to tune the XSA, and again the annex indicates they are. However, this is consistent with last years assessment procedures. These procedures should be clarified in the annex.
5 ) The annex should describe why 0-group survey indices are not used as a tuning index.
6 ) The year-classes mentioned in the first line of section 4.4.3 should be 20072009.

7 ) There is inconsistency in the numbers between Table 4.18 and Fig 4.1. Accordingly, Table 4.18 has been revised by the WG.
8 ) The prediction Table 4.20 indicate a SSB in 2009 of 285 kt . This is actually the SSB in start of the year 2010. Corrected by the WG.
9 ) Weights at age in the Norwegian survey has decreased for the oldest ages compared to last year, while the Russian survey shows decrease for all
ages. It is suggested to review data on weight at age matrices. There might be problems with the age reading presented by different nations. The Norwegian sampling program was terminated in Q3 2009 which might give less precise estimates of weight at age in the catch compared to previous years.
10 )The XSA is very sensitive to settings. There are several reasons for this: incomplete and variable between years survey coverage (both for Russian and Norwegian bottom trawl surveys), correlated error structures, biased catch statistics in relation to unknown discard and un-reported landings (IUU), predation on young age groups, and sampling error. The basis for this and key sources of this should be further investigated in a future benchmark assessment. The time series of un-reported landings was included this year as well, but the un-reported landings in 2009 was zero. Decreasing estimated IUU catches are explained in the Quality handbook. There are no estimates of discarding. Both Russian (2006) and Norw egian (2007) bottom trawl surveys coverage were reduced compared to previous years. There has been performed sensitivity analyses according to various XSA settings (Fig. 4.7).
11 )The swept area bottom trawl estimates from the joint Norwegian-Russian ecosystem survey have been tested as a new survey tuning time series in single fleet runs. The estimates from this were slightly lower compared to other single fleet runs. A run combining all time series was very close to the final XSA run, and inclusion of the new time series should be considered in a near future benchmark assessment.

12 )The assessment indicates that the increasing of the SSB is relative with decreasing F and due to the high level of recruitment. In general, there is consistency betw een different survey indices.
13 )The precautionary reference points are set based on an assessment carried out in 2000. The present assessment indicates that the historical biomasses estimates have been revised and that the technical basis for the biomass reference points is no longer valid. ICES needs to reconsider the MSY (and PA) reference points in a benchmark assessment in near future (2011).
14 )The technical review comments given to last years assessment has been addressed in this years working group report and assessment.
15 )There are different estimates of unreported catches/landings by Norway and Russian. As IUU catch estimates for 2009 is zero, the WG decided to make no comparisons and exploratory runs investigating the differences between assessments including each of the two time series. This years assessment only include the Norwegian data. As time series are still used with different perception of and assumptions associated to IUU it is recommended that these comparisons are still made in the assessment.
16 )There is a tendency that XSA estimates the peaks in abundance at age smoother than the surveys, which is consistent with aging error. This should be investigated in a near future benchmark assessment.
17 )Reference points were not revised due to time constraints by the WG, and this should be done at the next benchmark assessment.
18 )Retrospective runs for 2000-2002 show strong trends and look strange. Such a retro needs additional investigation in the next benchmark.

19 )Residuals for ages 7-8 for all surveys are high. There are also pronounced year effects in residuals. This should be investigated in the next benchmark assessment.

20 )An "index ratio by age" method was used to adjust for incomplete Joint Barents Sea winter survey coverage in some years (e.g. 2007). This should be revisited in time, because data before and after years with incomplete coverage can and should be used to fill in the missing data. It would also be desirable to reflect in survey standard errors the additional uncertainty caused by incomplete coverage. This may be more important in the future if more statistically rigorous state-space approaches are used, where process and measurement error are separated and it helps for this to have good information on within-survey error.
21 )Why are years not specified in Table 4.9A?
22 )The titles of Tables B1 and B3 should be changed to Joint Surveys.
23 )The annex contains insufficient detail in some aspects. For example, the annex and report are unclear if maturities are modelled by year or cohort?
24 )Section 4.3.6 is confusing. How has the same approach been used for predation and maturities? Are the changes to data important? If so, this should be described better.

25 )The annex does not specify the inputs to the RCT3 analyses. The procedures used were the same as last year.
26 )There are substantial differences in biomass and SSB in Table 4.18 of the this years report and Table 4.18 of last years report.
27 )Table 4.12 does not indicate if the XSA has converged.

## Conclusions

The assessment has been performed correctly as specified in the annex. There is need for a benchmark in the near future. The present management plan is in accordance with a precautionary approach and the stock is currently harvested sustainably. However, unreported catches and discards is an important issue for this stock and reduce the effect of management measures and the objectives of the harvest control rule. The information given by the assessments is sufficient to provide advice.

Stock: Saithe in Subareas I and II (Northeast Arctic) (report section 5)
Short description of the assessment: extremely useful for reference of ACOM!

1) Assessment type: update/SPALY from the February 2010 benchmark assessment (ICES CM 2010/ACOM:36).
2) Assessment: analytical
3) Forecast: analytical forecast presented
a) recruitment age 3 calculated as long term geometric mean (1960-2008);
b) stock numbers from XSA (age5+);
c) Pope's approximation used;
d) exploitation pattern is average of 2007-2009 (age 3-10) while average 20072009 for age 11-13 was used for age 11-15+;
e) three last years averages of w eight at age in the catch and stock used;
f) maturity average of 2008-2009.
g) $\mathrm{F}_{\text {status qou }}(2007-2009)=0.25$ and $\mathrm{F}_{\text {TAC } 204 \mathrm{kt}}=0.32$ are both below Fpa at 0.35 .
4) Assessment model: XSA is applied for the final assessment, using 4 tuning time series (2 acoustic survey cpue series (age 3-7) for the time span 1994-2001 and 20022009, and 2 trawl commercial cpue series (age 4-8, quarter 1-4) for the time span 19942001 and 2002-2009 according to benchmark recommendations - among other due to catchability shift in 2002);
a) maturity ogive - 3-year running average (1985-present);
b) $M$ fixed at 0.2 for all age groups;
c) annual estimates of weight at age in catch, and assumed catch at age in the stock to be equal to w eight at age in catch.
d) SSB is calculated at Jan 1st.
e) The settings of the XSA has been changed according to the February 2010 benchmark recommendations (see under Consistency below).
f) Discard is not included in the assessment.
g) Exploratory single fleet runs were performed among other with the new tuning time series from the combined survey but they did not perform as well as the "old" ones presently used mainly because of their short time span.
5) Consistency: Update assessment from the February 2010 benchmark assessment. Last year's assessment was accepted and used as a basis for advice in 2009.
a) Last years assessment estimated total stock (TSB (11+)) in 2008 to $4 \%$ higher and the SSB (11+) in $20084 \%$ lower than the previous assessment, while the present assessment estimates TSB (15+) 19\% lower, the SSB (15+) $34 \%$ lower, and the terminal year $\mathrm{F}(4-7) 25 \%$ higher in 2009 compared to previous years assessment (for TSB (11+), SSB (11+)).
6) Stock status: Stock is within safe biological limits. Fbar<Fpa and well below Flim; SSB>Bpa (since 1995) and well above Blim.
a) Recruitment is around average strength, the 2005 year class being slightly above long term mean, while the 2003 and 2004 year classes were poor.
b) New Fpa estimated in 2005 was accepted by ACFM.
c) The same biomass and fishing mortality reference points are used in this years assessment ( $\mathrm{Fpa}=0.35$, $\mathrm{Flim}=0.58$, $\mathrm{Bpa}=220 \mathrm{kt}$, and $\mathrm{Blim}=136 \mathrm{kt}$ ).
d) In the 2010 assessment, the age span was expanded from $11+$ to $15+$ and $\mathrm{im}-$ portant XSA parameters were changed due to benchmarking. Accordingly, reference points was re-estimated in 2010 using segmented regression. The results were a Bpa at 194 kt and a Fpa at 0.35.
e) The HCR is based on PA reference points, and if new ones are introduced, the HCR will have to be evaluated again. The WG explains that due to lack of time and transition to MSY based reference points the existing reference points was not changed. No attempts were made to set MSY reference points (FMSY and MSY Btrigger).
f) Fmax is estimated to 0.33 , F0.1 to 0.08 and $\mathrm{F}_{35 \%}$ SPR to 0.11 , and these points are Fmsy candidates, but the estimates (especially $\mathrm{F}_{\max }$ ) are unstable for this stock. Highest long-term yield was obtained for an exploitation level of 0.32.
g) The current $\mathrm{F}(0.27)$ is lower than the F associated with high long-term yield when applied within the agreed HCR.
7) Man. Plan.: There is a management plan for this stock, and a harvest control rule (HCR) is used for setting the annual TAC ( $\operatorname{target} \mathrm{F}_{\mathrm{pa}}$ ) which was in 2007 evaluated by ICES to be consistent with the precautionary approach. The implemented management plan implies a TAC based on the average catches for the coming 3 years based on $\mathrm{F}_{\mathrm{pa}}$ resulting in a TAC of 225 kt in 2009 and 204 kt in 2010 corresponding to $\mathrm{F}=0.29$ in 2009 and $\mathrm{F}=0.30$ in 2010. The fishery is regulated through TAC quota with max. $15 \%$ yearly change (when SSB is above $\mathrm{B}_{\mathrm{pa}}$ ). Landings in 2009 was 64 kt below the TAC. There are indications that the TAC will not be taken in 2010 according to the WG. In addition to TAC, the fishery is regulated through technical measures. Discarding, although illegal, occurs in the saithe fishery. There are no quantitative estimates of discard. Discard of young fish is by the WG not considered a major problem for the stock because they are inaccessible to commercial fishery due to their near shore distribution.

## General comments

This is a well documented, well ordered and well considered section. The assessment is due to the new benchmarking changes not fully consistent with last year's assessment, however, it is consistent with the benchmark assessment in February 2010. SSB is well above $\mathrm{B}_{\mathrm{pa}}$ and F is below $\mathrm{F}_{\mathrm{pa}}$, and the reference points are considered in accordance with the precautionary approach. The current F (0.27) is lower than the F associated with high long-term yield when applied within the agreed HCR.

## Technical comments

- The Stock Annex (Quality Handbook) has been revised. The review was restricted to a check whether the procedures described in the technical annex (handbook) were applied. This was evaluated to be the case. The benchmark assessment report (ICES CM 2010/ACOM:36) was not reviewed.
- The main conclusions from the recent benchmark assessment in February 2010 were: i) expand the catch matrix from 3-11+ to $3-15+$, ii) base the Norwegian trawl CPUE on data from all quarters from days with $>20 \%$ but $<80 \%$ saithe in catches, iii) split the two tuning series to before and after 2002, iv) reduce the shrinkage in the XSA (S.E. of the mean to which the estimate is shrunk increased from $0.5-1.5$ ) and remove time tapered down weighting. (ICES CM 2010/ACOM:36; this report was not reviewed).
- The Total Stock Biomass (TSB) for 2009 is not fully consistent between Table 5.5.5 (821055) compared to Table 5.5.7 (798292).
- The report refers to the retrospective patterns in Figures 5.8.1-2 which are illustrated in Fig. 5.5.5.
- Comparison with the last year report indicates better retrospective pattern obtained this year (more stable assessment) with the new benchmark settings than the one observed last year. The tendency to overestimate F and underestimate SSB in the terminal year seems to have changed to the opposite situation (see above). Despite these changes, the assessment is still evaluated to be in line with last year's assessment.
- There are still trends (age and year effects) in the residuals of the assessment. These tendencies should be explored and explained further in future assessments.
- Lack of reliable recruitment estimates is still a major problem. The survey recruitment indices are strongly dependent on the extent to which 2-4 year old saithe have migrated from the coastal areas and become available to the acoustic saithe survey on the banks, and this varies between years. The assessment and the forecast are sensitive to this, and the variability in this should be explored and discussed further.
- The biological sampling of some vessel groups may have become critically low after the termination of the Norwegian port sampling program in 2009. The effect of this should be explored further in future assessments.
- Medium term projections were made even though not considered reliable because the results are mainly driven by the assumption of mean recruitment and ignoring bias in the assessment. No improved recruitment estimates are available.
- Graphs of the tuning indices should be provided.
- Table 5.5.1 XSA run had not converged. This was considered in the 2009 report of WGMG.


## Remarks by the reviewer

-     - There is an indication in the sub-section "5.1 The Fishery" on saithe temporal substantial discarding occurring from non-Norwegian commercial trawlers. Although AFWG specifies that discarding is a minor problem, it could however be of some importance to investigate the level of discarding (by age) as this might have some impact on the perception of the stock dynamics (especially recruitment).
-     - Saithe has recently been more distributed southward and such was the biological sampling activity for estimating maturity ogives. Higher maturity rate in the southern area is observed. The 3 -year running average ogive used in the assessment is not weighted by abundance, and in consequence
it probably results in biased estimate of maturity ogive in the context of the whole stock.


## Conclusions

The assessment has been performed correctly. There has been performed benchmark assessment in 2010 which has improved the previous years problematic retrospective patterns. There are still trends in the residuals by age and year. The present assessment is consistent with the benchmark assessment. The information given by the assessments is sufficient to provide advice.

## Stock: Beaked redfish (Sebastes mentella) in Subareas I and II (report section 6)

Short description of the assessment: extremely useful for reference of ACOM!

1) Assessment type: update/SPALY
2) Assessment: no analytical assessment, survey trends.
3) Forecast: not presented.
4) Assessment model: not relevant.
5) Consistency: Survey data are still consistent and perception of stock status is unchanged.
6) Stock status: There are no reference points defined for this stock. All signals show that the stock has gradually declined and is at present near a low. Recruitment has failed since 1991, but the 2007 YC seems strong.
7) Man. Plan. There is no management plan. The fishery is managed with annual TAC and technical measures such as closed areas for certain gears. A description of the technical measures and history is given in section 2.1.2: Regulations.

## General comments

This was a well documented, well ordered and considered section. Most of the text was an update of last year's report but some new information on length composition of the demersal fishery (Sec. 6.2.1) and results from a new survey (Sec. 6.2.5.2) have been added. The description of the pelagic fishery has been moved to the stock annex.

## Technical comments

1 ) The review was restricted to a check whether the procedures described in the technical annex (handbook) were applied. The handbook was updated again this year.
2 ) Also, a comparison with last year's report was made.
3 ) No assessment was carried out and the WG restricted the work to updating tables. There were also changes in the text but most was revised from last year.
4 ) Why is Iceland in the Canada catch column of Table6.1, and Estonia in the Denmark column?

5 ) Is there a real need to have a second different set of tables and figures labeled starting with a D. These could be included with the other tables and figures, this would make reading easier.
6 ) In section 6.1.1 the reference to fig 6-2 should be deleted, same for figure 69 in 6.2.5.1, it doesn't deal with the ecosystem survey in August but with cod predation.
7 ) In section 6.1.3 the 2001 and 2006 landings figures don't match the values in the table 6.1.
8 ) Paragraph on cod's predation in section 6.2.5.1 should mention the data source.

9 ) What is the status of the new surveys described in 6.2.5.2 and 6.2.5.3? One time event or start of a series?

10 )Paragraph 6.3 reference to figure 6.4 is wrong, it is now figure 6.7 and although stated that there is an increase in fish $>30 \mathrm{~cm}$ in pelagic surveys, there is no data (table or figure) illustrating this.

## Remarks by the reviewer

- The unchanged perception of the stock compared to last year gives no reason to change previous advice.
- The report contains several chapters with information relevant to the advice
- The continued poor recruitment (decades), slow growth and late maturation gives no expectation that the stock will recover within the next 12-15 years. The only year classes that can contribute to the spawning stock in near future are those prior to 1991 as the following fifteen year classes are very poor. There are signs of increased recruitment at least in some areas (see figure 6.7 and 6.10) but the 2008 YC estimates are back to low value..
- Results from the pelagic surveys conducted in 2008 and 2009 indicate possible trawlable biomasses of about 400-500 kt but such estimates are highly imprecise. The section management advice in the report mention estimates of SSB from Anon (2009b) that are not discussed in the report and there not easily accessible (NEAFC WG report).
- There are no biological reference points defined for this stock. During the WKFRAME meeting, recommendations were made for this stock as it was used as a case study. The WG find it premature to adopt the values estimated by WKFRAME but will work on this prior to the benchmark assessment proposed for 2012


## Conclusions

There are no indications that there are changes in the stock status. The development of a fishery in international waters may be a source of concern, since the fishable stock consists of year classes before 1991 and there was poor recruitment thereafter. Traditional PA reference points may be not appropriate, but a more general approach on management advice could be adopted towards stocks with similar characteristics unless a gadget or other assessment method can be evaluated and adopted during the next benchmark assessment.

## Stock: Golden redfish (Sebastes marinus) in Subareas I and II (report section 7)

Short description of the assessment: extremely useful for reference of ACOM!

1) Assessment type: SALY
2) Assessment: assessed on the basis of available trends in the fisheries and surveys and an experimental analytical assessment.

The Gadget model was used for the sixth time as an experimental analytical assessment model.
3) Forecast: none.
4) Assessment model: model - tuning by 2 commercial fleets + 2 surveys (experimental)
5) Consistency: In this year's update the Gadget model configuration and settings were identical to that of 2009. Commercial catch data have been revised for 2008 and updated with year 2009. The general patterns in the stock dynamics are very similar to those modelled in 2008 but there is an increased discrepancy between the two surveys used in numbers at older ages and the years 2008 and 2009 have been excluded from the coastal survey.

## At ADGANW,

6) Stock status: The stock is currently in a very poor situation as confirmed by survey observations and Gadget assessment update. Reference points have not been defined.
7) Man. Plan. No Management Plan agreed.

## General comments

This was a well documented, well ordered and considered section. The text in the report is an update from last years' report. The tables and figures were unambiguous but the ordering in two sets (labeled $7 \ldots$ and D...) is questionable and makes the reading harder than necessary.

Note that in the state of the stock, there are a few paragraphs on the positioning of the species in the Norwegian Redlist as a vulnerable species and that following results from 2 workshops convened by ICES conclude that under current harvesting level and low recruitment there is a risk of stock collapse.

## Technical comments

1 ) The review was restricted to a check whether the procedures described in the technical annex (handbook) were applied. The handbook was updated again this year.
2 ) Also, a comparison with last year's report was made.
3 ) Tables numbering should be updated, there are no tables D1 etc but D11...
4 ) In 7.1.2. year range mentions 1983, while tables starts in 1989.
5 ) The total index in Figure 7.5a is somewhat different than the total index in Figure 7.5b, and there are discrepancies with Tables 12a,b. Indices in Fig $7.4 \mathrm{a}, \mathrm{b}$ are somewhat different as well. A description of the reasons for these differences would be helpful.
6 ) The data was used as specified in the stock annex.

7 ) The assessment experimental model been applied as specified in the stock annex but some changes have been made to the parameterisation of the Gadget model (revision of the 1998 data of the Barents sea survey, exclusion of the two most recent years from the Coastal survey and downweighting of the series in the fit).

8 ) The updated assessment gives a valid basis for advice as trends are overall similar.

9 ) However, there are some issues:
a) There is little information on the parameterization of the Gadget model and model fit except aggregated surveys residuals. The text table comparing the successive Gadget model's results shows a strong retrospective pattern (SSB in 2003 has been successively revised upward from 2006 to 2010 by a total of $44 \%$ ).
b) The fit is poor for the coastal survey but it has been down-w eighted in the fit. The overall fit seemed better in the 2009 formulation. There seems to be bias in the fits shown in Figure 7.6. Most points are above the lines. This could be an issue with convergence.
c) A constant selectivity through time was assumed in the model; the possibility of an extension with varying selectivity was mentioned by the group; this should be included in the next assessment.
d ) We had to go back to the 2007 report to find more on parameters estimated and likelihood components employed. And from this report: "The weighting of different components in a likelihood function is a clear problem in any model combining multiple data sources, and needs to be addressed in a wider fisheries assessment context in order for researchers to make best use of all the available data." The weightings of various data components should be described.
e) Although the assessment based on the Gadget model is experimental, it is a strong component of the advice on the state of the stock. There is a need to address better parameterization of the model used, including tuning indices selection now that there are increased discrepancies in the 2 surveys and a strong retrospective pattern. Why was the Inshore survey chosen and not the Norwegian Svalbard? There may be good reasons but it is not stated even in the Stock annex.
f) Retrospective plots (i.e. Fig. 7.8-7.10) should include assessments for several years.
g) the general problems of age reading in redfish should be addressed
h) Recruitment estimates may be biased due to species misidentification.

## Conclusions

The assessment has been performed correctly and gives a valid basis for advice as long as it is based on trends that are overall similar. A benchmark assessment is needed for this stock (expected in 2012) to address issues mentioned earlier. Until then, due to the expected low recruitment, the advice for this stock can be based on the assessment of the working group.

## Stock: Greenland halibut in subareas I and II (report section 8)

Short description of the assessment: extremely useful for reference of ACOM!

1) Assessment type: update/SPALY (update assessment in form of exploratory run used to view trends in the stock)
2) Assessment: analytical
3) Forecast: not presented
4) Assessment model: XSA (ages 5 and above) using same settings as last year and using 3 tuning fleets ( 2 surveys and 1 experimental commercial CPUE); estimates shrunk towards the mean of the final 2 years and 5 ages; S.E. of the mean to which the estimates were shrunk was set at 0.5; catch at age data and mean weigth at age in catch data for 2006-2009 are only available from Russian fisheries due to age determination problems; the mean weight at age in the stock is assumed equal to mean weight at age in catch; natural mortality set to 0.15 for all ages and years; a three year running aver age maturity ogive is used.
5) Consistency: The current assessment used the same catch matrix, survey series and settings as last year with updated data for 2008 and new data for 2009. Fishing mortalities tend to be overestimated while SSB tends to be underestimated, and recruitment tends to be overestimated in some recent years. The assessment is considered to be uncertain due to age-reading and survey data quality problems (e.g. uncertain recruitment and maturity indices as well as uncertain catch number and mean weight at age numbers). SSB is sensitive to the maturity ogive which is uncertain, and different survey indices shows opposite tendencies in proportion mature.
6) Stock status: The stock is currently stable at a relatively low level with a slightly increasing tendency. Fishing mortality in 2009 was at the same low level as in 2008. There are indications of low recruitment in the most recent year, while recruitment was high in the previous year. There are no reference points defined for this stock, and there is no estimate of high yield reference points. The catch in 2009 was 12 kt which is $6 \%$ less than the TAC and projected catch. The advice is that the catch in 2009 and 2010 should be below 13 kt , which is the level below which SSB has increased in the past. A TAC of 15 kt has been set for 2010 (see below) which is also the projected catch for this year by the WG. Discard and non-reported catches are not included in the assessment. Discard is not regarded to be significant and a problem by the WG, while it is believed that there may be some additional non-reported landings. The magnitude of the latter is not known.
7) Man. Plan.: No Management Plan agreed and no HCR set. There is no estimate of high yield reference points. The advice has not changed since 2003, yearly catches should be below 13000 t. This is the level below which SSB has increased in the past. The stock is regulated through by-catch regulations (max 12\% GHL in each haul and max $7 \%$ GHL in each landing, and annual catch of GHL should not exceed $4 \%$ of the sum of vessel quotas on cod, haddock and saithe and in total not exceed 40 t per vessel per year). Targeted Greenland halibut fishery has been forbidden in the period 1992 to 2009 except for a limited coastal Norwegian fishery where vessel specific TACs of 10-14 $t$ dependent on vessel size has been set. In 2009, it was decided to cancel the ban against directed Greenland halibut fishery, and a yearly TAC of 15 kt has been established for 2010-2012 shared between Norway ( $51 \%$ ), Russia ( $45 \%$ ) and other countries ( $4 \%$ ). There is allowed 4 kt catch by Norway and Russia each for research and surveillance purposes.

## General comments

This was a well ordered and well considered section. There might be some extensive detailing in the assessment report, and parts of this could with advantage be moved to the Stock Annex. The assessment and its results was easy to follow and to interpret. Due to age reading uncertainties as well as uncertainties in recruitment and maturity estimates, the stock assessment and advice is uncertain. The variability in these biological parameters has to be explored further in order to revise estimates for a future benchmark assessment.

## Technical comments

- In section 8.1.2 the years should be up-dated to 2009 and 2010.
- There are still retrospective patterns in the assessment. Fishing mortalities tend to be overestimated while SSB tends to be underestimated, and recruitment tends to be overestimated in some recent years.
- Exploratory runs of XSA have shown that there is high sensitivity in the assessment in relation the XSA parameter settings.
- Little is known about stock structure, stock delineation, distribution, and migration dynamics of the stock among other into other management areas. There is uncertainty concerning potential exchange between the Greenland halibut stock in the NEA and another stock in the Faeroe Is-lands-Iceland area and Greenland.
- The age structured tables of the Norwegian surveys have not been updated since 2006, due to change in age reading procedure as well as because of great problems and uncertainty in age reading. The new Norwegian age readings are not comparable with older data or the Russian age readings. This also influences estimates of recruitment and the maturity ogive significantly where the latter at present show very much variability. This needs to be considered and solved before a thorough benchmark assessment can be carried out. Age reading is addressed in the joint research program, and in a workshop in 2010. This will eventually end up in a total revision of the input data to the assessment. Russian agelength keys were used in the total catch matrix.
- It remains unknown which age should be used for a reliable recruitment estimate. The WG evaluates that shortcomings in estimation of recruitment is partly due to survey coverage. Future inclusion of northern parts of the Russian zone may improve the index.
- There are trends in catchability within individual surveys used for tuning of the XSA. Tuning time series for Norwegian surveys do not include the recent years from 2005 due to age reading problems.
- The assumption of $M=0.15$ needs to be explained. Additionally, the proportion of natural mortality before spawning is set to 0 . This also needs some explanation.
- Response to ACFM technical minutes: The technical review from last year has not been commented on by the WG.


## Conclusions

The ongoing age reading issue needs to be solved, and age reading revisions need to be completed before a reliable stock assessment can be performed. Age reading problems are the main concern for the assessment. There is an urgent need that this is
solved and consensus on age readings are reached - among other through the 2010 age reading workshop. This is needed among other to have reliable recruitment and maturity estimates as well as reliable catch number and weight at age matrices to be used in the assessment. Also, it is needed in order to include the more recent Norwegian survey tuning time series in the assessment.
In general there is a large uncertainty about the stock size so that conservative measures concerning fishing pressure on this stock are appropriate.

## Stock: Barents Sea Capelin (report section 9)

Short description of the assessment: extremely useful for reference of ACOM!

1) Assessment type: update (but no annex)
2) Assessment: Absolute survey estimates
3) Forecast: Probabilistic projection of the spawning stock to the time of spawning at 1 April 2010 was made using the spreadsheet model CapTool (implemented in the @RISK software).
4) Assessment model: none
5) Consistency: The assessment methodology appeared consistent with last years report.
6) Stock status: Blim=200 kt. Acoustic estimates of SSB for October in 2008 and 2009 are well above the Blim.
7) Man. Plan. The Mixed Russian-Norwegian Fishery Commission agreed to adopt a management strategy based on the rule that, with $95 \%$ probability, at least 200000 t of capelin should be allowed to spawn. With a catch of 360000 tonnes, the probability for the spawning stock in 2009 is below Blim is $5 \%$. The median spawning stock size in 2010 will then be 504000 tonnes. During its autumn 2009 meeting, the Joint Russian - Norwegian Fishery Commission decided that the quota according to the harvest control rule in 2010 will be 360000 tonnes

## General comments

There is no stock annex yet, and it was not possible to provide a technical review of this stock.

The acoustic survey estimates are treated as absolute. This requires defense. There is no assessment of the impact of past fishing on stock dynamics.

## Technical comments

None

## Conclusions

In October of 2008 and 2009 the stock was estimated to be well above the Blim.


[^0]:    * M output biomass is not calculated for 2009

[^1]:    * M output biomass is not calculated for 2009

[^2]:    * corrected data on cod consumption

[^3]:    ${ }^{1}$ Indices adjusted to account for limited area coverage.

[^4]:    Weights in kilograms

[^5]:    ${ }^{1}$ Provisional figures.
    ${ }^{2}$ Working Group figure.
    ${ }^{3}$ UK(E/W/)+UK(Scotl)

    + less than 0.5 ton.

[^6]:    ${ }^{1}$ - Includes some unidentified Sebastes specimens, mostly less than 15 cm .
    ${ }^{2}$ - Adjusted indices to account for not covering the Russian EEZ in Subarea I.

[^7]:    ${ }^{1}$ Provisional figures.
    ${ }^{2}$ Includes former GDR prior to 1991.
    ${ }^{3}$ USSR prior to 1991.
    4UK(E\&W)+UK(Scot.)

[^8]:    ${ }^{1}$ Provisional figures.
    2 Working Group figures.
    ${ }^{3}$ As reported to Norwegian authorities.
    ${ }^{4}$ USSR prior to 1991.

[^9]:    ${ }^{1}$ From 2004 part of the new joint ecosystem survey.
    Not updated from 2006 due to new age reading method

[^10]:    Not updated from 2006 due to new age reading method

[^11]:    *No survey in 2006-2007
    ** New swept area estimation method

[^12]:    ${ }^{1}$ Provisional figures

[^13]:    ${ }^{1}$ The average survey index in the years 1995-1998
    ${ }^{2}$ Ages 4-7

