# ICES WGNPBW 2005 

# Report of the Northern Pelagic and Blue Whiting Fisheries Working Group <br> (WGNPBW) 

25 August - 1 September 2005
ICES Headquarters Copenhagen

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Recommended format for purposes of citation:
ICES. 2005. Report of the Northern Pelagic and Blue Whiting Fisheries Working Group (WGNPBW), 25 August - 1 September 2005, ICES Headquarters Copenhagen. CM 2006/ACFM:05. 241 pp .

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The ICES northern pelagic and blue whiting fisheries working group (WGNPBW) met for 8 days in August 2005 to assess the state of two stocks, blue whiting and Norwegian spring spawning herring. Age-based assessments were carried out for both stocks. The assessment on the blue whithing stock was a full assessment as it is on the observation list while the assessment on the norwegian spring spawning herring stock was an update one. This year the assessments of the Icelandic capelin and the Icelandic summer spawning herring, which in previous year were by done by this working group, were done in spring in the NWWG.

For blue whithing 4 assessment models were used to explore the data. The exploration revealed conflict between catch data and surveys with models relying more on surveys estimating larger spawning stock in recent years. This year a number of recruitment indices were analysed, and the conclusion was that reasonable estimates could be obtained for the most recent yearclasses, in contrast to last year when geometric mean was used for the most recent yearclasses. The assessment was as in many recent years an upward revision of last year assessment.

The group expressed concern about high harvest levels in recent years and also that fishing mortality on the youngest age groups has been increasing in recent years despite the models failing to show this. The harvest levels will be far too high if recruitment will be reduced to earlier levels. Recruitment has been exceptionally good since 1996.

The assessment of Norwegian spring spawning herring was done by the same model as last year. Like last year, two alternating models gave quite different perception of the current stock size. This year's assessment is an upward revision of last year's assessment, but then, the last 5 assessments had been large downward revisions from earlier years. Some problems came up in estimating maturity at age and selection pattern of the big 2002 yearclass that has considerably different spatial distribution from all yearclasses in recent decades.

At ASC last year (2004) there was a proposal to merge the blue whiting and the norwegian spring spawning herring from NPBWWG and the mackerel and the horse mackerel from WGMHSA into one group. The idea behind this is that the fishery is in the same area for mackerel and blue whiting, which are both widely migrating species. The norwegian spring spawning herring is also widely migrating species, even though it does not migrate as south as the other species. All these species feed in the Norwegian Sea. To combine these species into one working group goes along with the regional approach that ICES has taken. The working group discussed this proposal. Mostly the response to this was positive. Some concerns were raised regarding the size of the group though.

The Stock Annexes for the Quality Control Handbook have not been made. They are left for the new working group.

## 1 Introduction

### 1.1 Participants

### 1.2 Terms of Reference

2ACFM07 The Northern Pelagic and Blue Whiting Fisheries Working Group [WGNPBW] (Chair: A. Gudmundsdottir, Iceland) will meet at ICES HQ from 25 August to 1 September 2005 to:
a) assess the status of and provide management options for 2006 for the Norwegian springspawning herring stock and the blue whiting stock;
b) provide as detailed information as possible on the age/size composition in different segments of the blue whiting fishery;
c) compile existing information on discards and by-catch by the fisheries;
d) enumerate the number, capacity and effort of vessels prosecuting the fishery by country.
e) for the stocks mentioned in a) perform the tasks described in C.Res. 2ACFM01.

WGNPBW will report by 2 September 2005 for the attention of ACFM.
In ToR e) referring to C.Res.2ACFM01 is given below:
WGNSSK, WGSSDS, WGHMM, WGMHSA, WGBFAS, WGNSDS, WGNPBW, AFWG, HAWG, NWWG, and WGPAND will, in addition to the tasks listed by individual group, in 2005:
(1) for stocks where it is considered relevant, review limit reference points (and come forward with new ones where none exist) and develop proposals for management strategies including target reference points if management has not already agreed strategies or target reference points (or HCRs) - following the guidelines from SGMAS (2005) and AMAWGC (2004 and 2005);
(2) comment on the outcome of existing management measures including technical measures, TACs, effort control and management plans;
(3) based on input from WGRED incorporate (where appropriate) existing knowledge on important environmental drivers for stock productivity and management into assessment and prediction, and important impacts of fisheries on the ecosystem;
(4) update the description of fisheries exploiting the stocks, including major regulatory changes and their potential effects. The description of the fisheries should include an enumeration of the number, capacity and effort of vessels prosecuting the fishery by country;
(5) where misreporting is considered significant provide information on its distribution on fisheries and the methods used to obtain the information
(6) provide for each stock information on discards (its distribution in time and space) and the method used to obtain it. Describe how it has been considered in the assessment;
(7) provide on a national basis an overview of the sampling of the basic assessment data for the stocks considered;
(8) provide specific information on possible deficiencies in the 2005 assessments including, at least, any major inadequacies in the data on landings, effort or discards;
any major inadequacies in research vessel surveys data, and any major difficulties in model formulation; including inadequacies in available software. The consequences of these deficiencies for both the assessment of the status of the stocks and the projection should be clarified.

### 1.3 Methods/software used

Through the years this group has used some softwares for data exploration and assessment of the stocks both separable models and a VPA-type models. These models are: AMCI, ICA, ISVPA, SMS and SeaStar.

### 1.3.1 AMCI

The assessment model AMCI (Assessment Model Combining Information from various sources), version 2.1, was described in the Working Group report in 2002. For assessments in 2003 AMCI version 2.2 (May 2002) has been used. This version is essentially an updated version of AMCI 2.1 where some known problems have been solved but without important changes in functionality. An updated manual was available for the Working Group. The Working Group on Methods on Fish Stock Assessments explored and evaluated AMCI 2.2, together with ISVPA and an array of other assessment models in their meeting in early 2003 (ICES 2003/D:03). The report of that Working Group can be consulted for more details on AMCI. For the assessment in 2004 a new version was used, AMCI 2.3a. A new version of AMCI, version 2.4 was available for the Working Group and was used.

### 1.3.2 ICA

ICA (Patterson, 1998) is a separable model over a recent number of years and a conventional VPA over the earlier part of the time series.

### 1.3.3 ISVPA

ISVPA version is basically the same as was used last year, except that now it can account for surveys in the termina+1 year, and is described in the table below.

| Model | ISVPA |
| :--- | :--- |
| Version | $\mathbf{2 0 0 4 . 3}$ |
| Model type | A separable model is applied to one or two periods, determined by the user. The <br> separable model covers the whole assessment period |
| Selection | The selection at oldest age is equal to that of previous age; selections are <br> normalized by their sum to 1. For the plus group the same mortality as for the <br> oldest true age. No manned inputs. |
| Estimated <br> parameters | The catchabilities by ages and fleets can be estimated or assumed equal to 1. <br> Catchabilities are derived analytically as exponents of the average logarithmic <br> residuals between the catch-derived and the survey-derived estimates of <br> abundance. |
| Catchabilities |  |


| (terminal + 1) | (terminal +1 ) is equal to that of terminal year) |
| :---: | :---: |
| Objective function | The objective function is a weighted sum of terms (weights may be given by user). For the catch-at-age part of the model, the respective term is: <br> - sum of squared residuals in logarithmic catches (SS), or <br> - median of distribution of squared residuals in logarithmic catches MDN(M, fn ), or <br> - the median of the distribution of absolute deviations of residuals from their median value - absolute median deviation $\operatorname{AMD}(\mathrm{M}, \mathrm{fn})$. <br> For SSB surveys it is sum of squared residuals between logarithms of SSB from cohort part and from surveys. <br> For age- structured surveys it is SS, or MDN, or AMD for logarithms of $\mathrm{N}(\mathrm{a}, \mathrm{y})$ or for logarithms of proportions-at-age, or for logarithms of weighted (by abundance) proportions-at-age. |
| Variance estimates/ uncertainty | For estimation of uncertainty parametric conditional bootstrap with respect to catch-at-age, (assuming that errors in catch-at-age data are log-normally distributed, standard deviation is estimated in basic run), combined with adding noising to indexes (assuming that errors in indexes are log-normally distributed with specified values of standard deviation) is used. |
| Other issues | Three error models are available for the catch-at-age part of the model: <br> - errors attributed to the catch-at-age data. This is a strictly separable model ("effort-controlled version") <br> - errors attributed to the separable model of fishing mortality. This is effectively a VPA but uses the separable model to arrive at terminal fishing mortalities ("catch-controlled version") <br> - errors attributed to both ("mixed version"). For each age and year, F is calculated from the separable model and from the VPA type approach (using Pope's approximation). The final estimate is an average between the two where the weighting is decided by the user or by the squared residual in that point. <br> Four options are available for constraining the residuals on the catches: <br> 1. Each row-sum and column-sum of the deviations between fishing mortalities derived from the separable model and derived from the VPA-type model are forced to be zero. This is called "unbiased separabilization" <br> 2. As option 1, but applied to logarithmic catch residuals. <br> 3. As option 1, but the deviations are weighted by the selection-at-age. <br> 4. No constraints on column-sums or row-sums of residuals. |
| Program language | Visual Basic |
| References | Kizner Z.I. and D.A.Vasilyev. 1997. Instantaneous Separable VPA (ISVPA). ICES Journal of Marine Science, 54, N 3: 399-411 <br> Vasilyev, D.A. (2001). Cohort models and analysis of commercial bioresources at information supply deficit. VNIRO Publishing: Moscow. <br> Vasilyev D. 2003. Is it possible to diminish the impact of unaccounted time trends in age structured surveys' catchability on the results of stock assessment by means of separable cohort models ? ICES CM 2003/X:03. 13 pp. <br> Vasilyev, D. 2004a. Winsorization: does it help in cohort models? ICES CM 2004/K:45. 19 p. <br> Vasilyev, D. 2004. Description of the ISVPA (version 2004.3) WP to ICES WGMHSA (2004), 24 p. |

### 1.3.4 SeaStar

SeaStar model is documented on the web site www.assessment.imr.no, where the user guide and the Mathematica code can be found, as well as supplementary documentation material. Also, a pdf file of the documentation is available from the author and at ICES.

SeaStar is statistically based. All terms in the likelihood function express the probability for the observation, where the expectation value is given by the modelled stock and the variance depends on parameters that are estimated together with the other tuning parameters. This avoids some of the subjectivity often found in assessments, where specific weights must be given to the various time series of data used for tuning. However, subjective choices must be made at various points in the assessment, for instance which tuning data that are to be used or whether outliers should be excluded.

### 1.3.4.1 Tuning

SeaStar is a traditional back-calculating tuning model using a VPA based on Pope’s approximation. If needed, solving the catch equation is implemented in case the model should be used for a stock with high fishing mortality. The stock is assessed by running the VPA, which is dependent on the F-values in the last year and the F-values for the oldest true age group. Taking the historic stock as the expectation value in underlying distributions for the observed survey data the probability of observing the survey data is calculated and included in the likelihood function. There is provision for selecting different functions to describe the survey distribution. In the present tuning the gamma distribution with a constant CV is used, in accordance with recent practice. Similarly, the probability of observing the tag return data is calculated and included in the likelihood function. It is assumed that the probability of tag returns, which are rare events, follows a Poisson distribution. At the 2000 meeting also a larval observation series was added, where the probability of observation is based on the spawning stock.

The historic stock is assessed by varying the unknown parameters until the maximum of the likelihood function is reached. The parameters that usually are varied (free parameters, tuning parameters) are:

Catchabilities for the tuning data
Uncertainty parameters for the data
Tagging survival
Terminal F-values
SeaStar provides for basing the likelihood only on the strongest year classes (here referred to as tuned year classes). Also, only the terminal F values for the strongest year classes may be used as tuning variables. The rationale for this is to stabilize the tuning by avoiding bias from large relative errors in the catch in the terminal year of weak year classes, which mediated by the catchabilities would propagate also to the stronger year classes. The terminal F values of the weak year classes are linearly interpolated between the terminal F values that are tuning parameters. The terminal F values of the fish younger than the youngest tuned year class is linearly interpolated to zero at age -1 . However, in recent years the practice has been to include all year classes that reside in the Barents Sea in the tuning, so that the latter feature is of no consequence for the assessment. The choice of tuned year classes is subjective. Of importance is to avoid the weakest year classes. The Norwegian spring spawning herring has extremely dynamic recruitment and most often the choice is rather obvious.

In SeaStar it is possible to perform the estimation in separate steps, where during one step the parameters estimated in the previous steps may be used, and new parameters can be estimated.

This feature is used for Norwegian spring spawning herring in order to first fit the adult part of the stock to the main tuning series in the Norwegian Sea and along the Norwegian coast. Keeping this part of the stock fixed the young part of the stock is estimated using survey series from the Barents Sea, which are considered more uncertain. The advantage is that the uncertain data from the Barents Sea then will not influence the estimate of the adult stock, that is of the larger importance in the short term. The adult year classes at ages 1 and 2 enter into the likelihood terms for the Barents Sea data, though, in order to provide a better basis for the estimation of catchabilities for those data.

The most important output variable is the estimated spawning stock in the assessment year, which is calculated on the basis of number at age, weight at age and maturity at age at January 1 in the assessment year. Number at age is taken from the VPA by calculating forward one year using the catch information in the last year. Weight at age and proportion mature at age in the assessment year are input data. However, it is assumed that the spawning occurs timeBeforeSpawning part of a year into the assessment year and in order to calculate the decrease until spawning time the same F as in the last year of catch is assumed also to apply for the assessment year. However, for the short term projection the WG also will assume that a fixed catch of catchAssessmentYear million tonnes will be taken in the assessment year, which may correspond to a somewhat different F .

The number of fish in the plus group is calculated as the sum of the number of fish in the oldest true age group and in the plus group the year before, reduced by natural mortality and catch. By applying the customary natural mortality of 0.15 also for the plus group, the smallest spawning stock biomass in the 1970s was about 0.5 million tonnes, which has been perceived as far to much concerning the collapsed state of the herring at that time. Thus, this year SeaStar was provided with the option of an increased natural mortality in the plus group. By setting this to 0.5 the smallest spawning stock in the 1970s is acceptable small. This did not lead to any appreciable reduction of present spawning stock biomass.

### 1.3.4.2 Catchabilities

When SeaStar is used for tuning Norwegian spring spawning herring, flat catchabilities (scalars) are used for fish that have recruited to the survey, i.e. fish that are older than a threshold age specified for each survey. However, there are provisions for making the catchabilities dependent on age as well as on cohorts.

### 1.3.4.3 Calculation of the plus group

In SeaStar, the numbers in the plus group is calculated as the number in the plus group the year before plus the number in the oldest true age group the year before, adjusted for M in the plus group and catch the year before. Thus, no assumption about the F-value in the plus group is needed. However, with low M-value on the plus group, the number of fish in the plus group may accumulate to unrealistic levels.

### 1.3.4.4 Analysis of assessment uncertainty using bootstrap

The analysis of assessment uncertainty is done using bootstrap. The assessment is run many times, each time new data sets are generated by resampling from the original data set.

### 1.3.4.5 Surveys

The surveys are resampled from the distribution that is assumed when the likelihood function is constructed, based on the unperturbed surveys.

### 1.3.4.6 Tagging

The number of tags recovered are sampled from the same distribution as assumed when the likelihood function is evaluated, i.e. Poisson. The number of fish screened for tags is assumed normally distributed with a CV specified in the input data list. The uncertainty in the number of screened fish stems from uncertainty regarding the amount of fish screened and uncertainty in the calculation of number at age screened form biological samples taken from the catch. The number of tagged fish released is also assumed uncertain where a normal distribution with a CV specified in the input data list is assumed.

### 1.3.4.7 Catch

The catches are considered certain so there is no distribution from which to draw catch data. The best method would be to base the catch data bootstrap on the biological samples used for distributing the catch on age. However, a possibly large source of error in the age distribution of the catch data comes from using biological samples from one space-time domain on catches from another space-time domain. This is necessary because of inadequate biological sampling of the catch from the countries involved in the fishery. The associated error cannot be dealt with however without implementation of the biological samples from all countries and using a time-space model of the fish distribution. This is an important but large project that ideally should be a joint effort of the countries involved.

When SeaStar is used for tuning Norwegian spring spawning herring it is assumed that the error is catch stems from misreading the age by one year. For two neighbouring age groups with number at age of stock1 and stock2 (as based on the unperturbed assessment) the catch to transfer is calculated as:
transferred $=\max \operatorname{Transfer} \frac{\text { Abs }(\text { stock } 1-\text { stock } 2)}{\text { stock } 1+\text { stock } 2}$
where maxTransfer is a setting.

### 1.3.4.8 Larvae

The larval data are resampled from the assumed distribution.

### 1.3.5 SMS

SMS (Stochastic Multi Species model; Lewy and Vinther, 2004) is an age-structured multispecies assessment model which includes biological interactions. However, the model can be used with one species only. In "single species mode" the model can be fitted to observations of catch-at-age and survey CPUE. SMS uses maximum likelihood to weight the various data sources assuming a log-normal error distribution for both data sources. The likelihood for the catch observation is then as defined below:

$$
L_{C}=\prod_{a, y, q} \frac{1}{\sigma_{\text {catch }}(a a) \sqrt{2 \pi}} \exp \left((\ln (C(a, y, q))-\ln (\hat{C}(a, y, q)))^{2} /\left(2 \sigma_{\text {catch }}^{2}(a a)\right)\right)
$$

where $C$ is the observed catch-at-age number, $\hat{C}$ is expected catch-at-age number, $y$ is year, $q$ is quarter, $a$ is age group, and $a a$ is one or more age groups.

SMS is a "traditional" forward running assessment model where the expected catch is calculated from the catch equation and $F$-at-age, which is assumed to be separable into an age selection, a year effect and a season (year, half-year, quarter) effect.

As an example, the F model configuration is shown below for a species where the assessment includes ages $0-3+$ and quarterly catch data and quarterly time step are used:
$F=F\left(a_{a}\right) \times F\left(y_{y}\right) \times F\left(q_{q}\right)$,
with $F$-components defined as follows:
$F(a)$ :

| Age 0 | $\mathrm{Fa}_{0}$ |
| :--- | :--- |
| Age 1 | $\mathrm{Fa}_{1}$ |
| Age 2 | $\mathrm{Fa}_{2}$ |
| Age 3 | $\mathrm{Fa}_{3}$ |

$F(q):$

|  | q 1 | q 2 | q 3 | q 4 |
| :--- | :--- | :--- | :--- | :--- |
| Age 0 | 0.0 | 0.0 | Fq | 0.25 |
| Age 1 | $\mathrm{Fq}_{1,1}$ | $\mathrm{Fq}_{1,2}$ | $\mathrm{Fq}_{1,3}$ | 0.25 |
| Age 2 | $\mathrm{Fq}_{2,1}$ | $\mathrm{Fq}_{2,2}$ | $\mathrm{Fq}_{1,3}$ | 0.25 |
| Age 3 | $\mathrm{Fq}_{3,1}$ | $\mathrm{Fq}_{3,2}$ | $\mathrm{Fq}_{3,3}$ | 0.25 |

$F(y)$ :

| Y 1 | Y 2 | Y 3 | Y 4 | Y 5 | Y 6 | Y 7 | Y 8 | Y 9 | $\ldots$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | $\mathrm{Fy}_{2}$ | $\mathrm{Fy}_{3}$ | $\mathrm{Fy}_{4}$ | $\mathrm{Fy}_{5}$ | $\mathrm{Fy}_{6}$ | $\mathrm{Fy}_{7}$ | $\mathrm{Fy}_{8}$ | $\mathrm{Fy}_{9}$ | $\ldots$ |

The parameters $F\left(a_{a}\right), F\left(y_{y}\right)$ and $F\left(q_{q}\right)$ are estimated in the model. $F\left(q_{q}\right)$ in the last quarter and $F\left(y_{y}\right)$ Fy in the first year are set to constants to obtain a unique solution. For annual data, the $F\left(q_{q}\right)$ is set to a constant 1and the model uses annual time steps.

One $F(a)$ vector can be estimated for the whole assessment period, or alternatively, individual $F(a)$ vectors can be estimated for subsets of the assessment periods. A separate $F(q)$ matrix is estimated for each $F(a)$ vector.

For the CPUE time series the expected CPUE numbers are calculated as the product of an assumed age (or age group) dependent catchability and the mean stock number in the survey period.

The likelihood for CPUE observations, $L_{S}$, is similar to $L_{C}$, as both are assumed lognormal distributed. The total likelihood is the product of the likelihood of the catch and the likelihood for CPUE ( $L=L_{C} * L_{\text {CPUE }}$ ). Parameters are estimated from a minimisation of $-\log (L)$.

The estimated model parameters include stock numbers the first year, recruitment in the remaining years, age selection pattern, and the year and season effect for the separable $F$ model, and catchability at age for CPUE time series.

SMS is implemented using the Ad-model builder (Otter Research Ltd.), which is a software package to develop non-linear statistical models. The SMS model is still under development, but has extensively been tested in the last year on both simulated and real data.

SMS can estimate the variance of parameters and derived values like average $F$ or SSB from the Hessian matrix. Alternatively, variance can be estimated by using the built-in functionality of the AD-Model builder package to carry out Markov Chain Monte Carlo simulations (Gilks et al. 1996), MCMC, to estimate the posterior distributions of the parameters. For the historical assessment, period uniform priors are used. For prediction, an additional stock/recruitment relation including CV can be used.

## 2 Ecological considerations

### 2.1 Ecosystem overview

### 2.1.1 Barents Sea

An overview of the ecological status of the Barents Sea in 2005 is given by the AFWG (ICES 2005).

### 2.1.2 Norwegian Sea

### 2.1.2.1 Hydrography and climate

The Nordic Seas (Fig. 2.1.2.1.1) during the last decades have been characterized by increased input of Arctic waters. The Arctic waters to the Norwegian Sea are mainly carried by the East Icelandic Current (EIC) and also to some extent by the Jan Mayen Current. During periods of increased Arctic water input, the western extension of Atlantic water is moved eastward. As a result, over the last 25 years the southern and western Norwegian Sea has become colder and fresher while the eastern Norwegian Sea is warmed. Atmospheric forcing drives this trend. From mid 1960's the winter North Atlantic Oscillation index (NAO) has increased to beginning of the 1990s followed by a reduction to the long-term-mean (Fig. 2.1.2.1.2). NAO is an indicator of the strength of the westerly winds into the Norwegian Sea. A high NAO index (i.e. stronger westerly winds) will force Atlantic and Arctic waters more eastward. In winter 2005 the index was about normal. However, a closer look into the monthly values shows that the NAO index for the winter 2005 was relatively high in both December and January but low in February and March. The high values in December-January during winter 2005 can then explain the more eastward displacement of the EIC in 2005 compared to 2004 (see below).

The Institute of Marine Research, Norway, has measured temperature and salinity in three standard sections in the Norwegian Sea almost regularly since 1978. The sections are 1) the Svinøy section which runs NW from $62.37^{\circ} \mathrm{N}$ at the Norwegian coast, 2) the Gimsøy section which also runs NW from the Lofoten Islands and 3) the Sørkapp section which is a zonal section at $76.33^{\circ} \mathrm{N}$ just south of Svalbard.

Figure 2.1.2.1.2 shows the development in temperature and salinity in three different sections from south to north in the Norwegian Sea. During the last 10 years the temperature and salinity in the Svinøy section have increased linearly. The temperature was in 2005 above normal but still less than in 2002-2004 that were the three warmest years in the time series. In 2005 the salinity in the Svinøy section had the largest value in the time series, about 0.09 above normal. Unfortunately some data are missing in the Gimsøy and Sørkapp sections during the last years. In 2004 for both sections the temperature and salinity were above the long-term-means.

The area of Atlantic water (defined with $\mathrm{S}>35.0$ ) in the Svinøy-section has been calculated. The mean temperature within the limited area has also been calculated, and the results for both spring and summer are shown in Fig. 2.1.2.1.3. There are considerable variations both in the area of Atlantic water distribution and its temperature. The temperature has shown a steady increase and since 1978 the Atlantic water has been about $0.7^{\circ} \mathrm{C}$ warmer (linearly). During the years 1992-1995 the area was much smaller than average for both seasons, probably due to strong westerly winds. The summer temperature had the three largest values in the time series during 2002-2004 but in 2005 it dropped, close to the long-term-mean. The area of Atlantic water was in 2005 above, but close to, the long-term-mean for both spring and summer.

During research cruises in May with the aim of measuring the stock size of pelagic fishes, hydrographic observations are also taken, covering most of the Norwegian Sea (figures are not shown). In May 2005 there was a larger influence of Arctic water in the southern Norwegian Sea compared to 2004, probably due to stronger westerly winds in December-January compared to last year. In the eastern part of the Norwegian Sea it was reported that the Atlantic Water was about $0.5-1.0^{\circ} \mathrm{C}$ colder in 2005 than in 2004. A research cruise was also performed in the northern Norwegian Sea during August 2005. Compared to 2003 the Atlantic water in August 2005 was considerable warmer (about $1^{\circ} \mathrm{C}$ ).

## Conclusions:

- The winter NAO index in 2005 was close to but still lower than normal. However, the index was relatively large in December-January.
- In 2005, there was an increased influence of Arctic water, from the EIC, in the southern Norwegian Sea compared to 2004.
- In the eastern Norwegian Sea the Atlantic Water was about $0.5-1.0^{\circ} \mathrm{C}$ colder in May 2005 compared to May 2004.
- In August 2005, the salinity in the core of Atlantic Water (at the slope, near the shelf) in the Svinøy section was record high, about 0,09 above the long-term-mean.
- The summer temperatures of Atlantic Water in the Svinøy section have been the highest ever during 2002-2004 but in 2005 it dropped but was still above the long-term-mean.
- The averaged summer temperature of the Atlantic Water in the Svinøy section has increased linearly with approximately $0.7^{\circ} \mathrm{C}$ since 1978 .


### 2.1.2.2 Phytoplankton

The development of phytoplankton in the Atlantic water is closely related to the increase of incoming solar irradiance during March and to the development of stratification in the upper mixed layer due to warming. The Institute of Marine Research, Norway, started in 1990 a long-term study of the mechanisms controlling the development of phytoplankton at Ocean Weather Station Mike situated at $66^{\circ} \mathrm{N}, 2^{\circ} \mathrm{E}$. It was not possible to get data from 2004 and 2005 ready for this report.

Figure 2.1.2.2.1 shows the development of the phytoplankton bloom for 2003 expressed as chlorophyll a concentration at the surface. In previous years there has been a marked difference in the time when the spring bloom reached its maximum. In 1997 the spring bloom reached its maximum 20 May (day of the year 140), in 1998 about one month earlier 18 April (day of the year 108). The timing of the bloom in 1999 was similar to that in 1998, but did not show the same high maximum in chlorophyll. This may be related to the weekly measurements in 1999, as opposed to daily measurements in 1997 and 1998. On the other hand, weekly measurements prior to 1997 have revealed pronounced maxima in chlorophyll. The reason for the low algal biomass in 1999 may have been early and strong grazing from a large over-wintered zooplankton stock. In all these years a strong peak has characterized the bloom. The situation in 2001 was different to previous years. First, the spring bloom started somewhat later (first week of May) compared to 1998 and 1999 and was followed by relatively moderate chlorophyll concentrations culminating with a major peak in the first week of June. Also a distinct early autumn bloom was observed in the middle of August. In 2002 the springbloom started to develop in the middle of April reaching its maximum at the end of April, resulting in one of the earliest bloom second only to the bloom in 1998. The 2003 bloom also maintained relatively high chlorophyll concentrations for about a month after the first peak on May 8 to decrease rapidly afterwards. After the main spring bloom four other smaller blooms were observed throughout the summer and early autumn.

The development of the phytoplankton prior to the spring bloom may be separated into two phases. The first phase, from day 1 to about day 50, is characterised by extremely low phytoplankton biomass expressed as chlorophyll $a$. This is the winter season during which phytoplankton growth is mainly limited by the low incoming irradiance typical of this period. The second phase, from about day 50 to day 100, is characterised by a gradual increase of phytoplankton biomass but without reaching bloom conditions. This is the pre-bloom phase during which the increase in biomass is related to the increase in incoming irradiance and the lack of a bloom is due to the deep upper mixed layer still present at this time.

Figure 2.1.2.2.2 shows the extension in time for these two phases and the timing of the spring bloom for the period 1991-2003. In a "normal" year the winter season extends to about 2 March. The pre-bloom phase extends on average from the 2 March to 16 April. The spring bloom starts normally on 16 April and reaches its maximum on 21 May, but the year-to-year variations are much larger than those of the previous phases. From 1991 to 1995 the trend was towards earlier spring blooms. This trend was broken in 1996, and thereafter year-to-year variability in the timing of the bloom has been greater but with a trend towards earlier blooms again after 2001.

Conclusions:

- The phytoplankton bloom in 2003 developed earlier than the average since 1991, third only to the 1998 and 2002 blooms.
- Chlorophyll $a$ concentrations first peaked in the first week of May 2003 and were maintained at relatively high levels until the first week of June resulting in the longest bloom in the time series. This could, as in 2002, have been the result of a relaxation in the grazing pressure.
- During summer and early autumn of 2003 several peaks of relatively high chlorophyll $a$ concentration were observed indicating a strong variability in minor blooms.


### 2.1.2.3 Zooplankton

Zooplankton biomass distribution in the Norwegian Sea has been mapped annually in May (since 1995) and in July (1994-2002). The sampling in July probably will be resumed in 2006. Zooplankton samples for biomass estimation were collected by vertical net hauls (WP2) or oblique net hauls (MOCNESS). In the present report zooplankton samples from the upper 200 m are analysed. Total zooplankton biomass ( g dry weight $\mathrm{m}^{-2}$ ) in May was averaged over sampling stations within three water masses, Atlantic water (defined by salinity $>35$ at 20 m depths), Arctic water (salinity $<35$, west of $1.4^{\circ} \mathrm{E}$ ) and Coastal water (salinity $<35$, east of $1.4^{\circ} \mathrm{E}$ ) (Fig. 2.1.2.3.1). In Atlantic and Arctic water masses zooplankton biomass decreased to a minimum in 1997. Thereafter zooplankton biomass increased again and remained relatively high except for a temporary reduction in 2001. After 2002 there has been a continuous reduction and in 2005 the second lowest biomass during the time series was measured. Due to reduced cruise time the Arctic water mass was not sampled in 2001 and 2004. Zooplankton biomass in Arctic water is generally higher than in Atlantic and coastal water, but in 2002 and 2005 the biomass in Arctic and Atlantic water equalled. In 2005 the highest biomass of the Norwegian Sea was found in coastal water. In the coastal water mass, which includes the Norwegian continental shelf and slope waters influenced by Norwegian coastal water, the temporal pattern of variation in biomass is different from the other two water masses.

In July the total zooplankton biomass ( g dry weight $\mathrm{m}^{-2}$ ) in the upper 200 m was calculated by integrating biomass at sampling stations within a selected area in the central and eastern Norwegian Sea. There is no obvious trend in the zooplankton biomass in July since 1994 (Figure 2.1.2.3.2).

Conclusions:

- Average zooplankton biomass in Atlantic water masses of the Norwegian Sea in May 2005 was much lower than average and the second lowest for the time series.
- Biomass in coastal water in 2005 was for the first time higher than in Arctic and Atlantic water.


### 2.1.2.4 Predictions for zooplankton biomass

The North Atlantic Oscillation index (NAO), is a proxy for the strength and duration of southwesterly winds, and is correlated with the inflow of Atlantic water to the Norwegian Sea. In the Norwegian Sea the average biomass of zooplankton in Atlantic water in May is fairly well correlated with the average NAO for the March-April period the previous year (Fig. 2.1.2.4.1). However, the model has consistently overestimated the biomass since 2003. This may be related to changes in the processes underlying the relationship, however, changes in for example the area covered by the Norwegian vessel during the international Norwegian Sea survey is another possible reason. March-April is the period when the primary production in the Norwegian Sea is initiated and the major reproductive period for many important zooplankton species such as Calanus finmarchicus and krill. The one-year lag in the relationship may be because we in May mainly measure the size of the overwintering stock, i.e. the previous years production and the present years spawning stock. Based on this relationship the biomass for May 2006 is estimated at 10.9 g dry weight $\mathrm{m}^{-2}$. Due to the tendency towards overestimation by the model during the last years, we perceive this as an overestimate as well.

Biomass (yr2) $=2.23 * N A O$ yr1 +10.54
$R^{2}=0.47, P=0.02$
Conclusions:

- The average NAO for March-April the previous year is directly related to zooplankton biomass in May and herring condition in the autumn.
- The biomass of zooplankton in 2006 is predicted at 10.9 g dry weight $\mathrm{m}^{-2}$ by the model, but is expected to be somewhat lower.


### 2.1.3 Icelandic waters

### 2.1.3.1 Hydrography and climate

As Iceland is situated at a meeting place of warm and cold currents its waters are characterised by highly variable conditions especially in the area north and northeast of the country. Heat and salt content in those waters depend on the strength of Atlantic inflow through the Denmark Strait and the variable flow of polar water from the north with the East Icelandic Current. South and west of Iceland fluctuations are smaller.

Climatic conditions in the North Atlantic improved around 1920 and remained rather warm until the mid-1960s, when they deteriorated. In the area north and east of Iceland temperature and salinity declined sharply in 1965 and these severely cold conditions lasted until 1971. After that, climatic conditions off north and east of Iceland improved, but were variable and years have alternated with cold years (Fig. 2.1.3.1.1).

Salinity and temperature increased in 1997 west of Iceland and have remained high. This increase in the Atlantic character of the Irminger Current reached into the northern area and peaked in 1999 and prevailed until winter and spring 2002 when a rather short period of polar influence was observed. In summer 2002 a persistent inflow of Atlantic water started and kept
on throughout the year 2003 with little winter cooling which resulted in a record year in temperature and salinity north and east of Iceland. Observations in February 2004 showed continued influence of Atlantic water in the northern and eastern area with stronger winter cooling and high salinities.

Temperature in the warm Atlantic waters south and west of Iceland in 2005 was $5-7^{\circ} \mathrm{C}$ and salinity $35.10-35.27$, i.e. high values like during the last years. There was a considerable Atlantic inflow eastwards onto the N-Icelandic shelf, reaching to the east of Melrakkasletta. However, there was a fresh and moderately warm surface layer over most of the Atlantic water, north, northeast and east of Iceland - remains of the ice which drifted east off the north coast earlier this spring. Nevertheless, both temperature and salinity in the upper layers were around or higher than the long term average north of Iceland, but somewhat lower than during most recent years.

In the East Icelandic Current outside the shelf edge northeast of Iceland, temperature and salinity were near the long-time average, while the southern limit of the 'cold tongue' reached further south than during the last years. East of Iceland, temperature and salinity in the upper layers were $2-3^{\circ} \mathrm{C}$ and $34.5-34.9$ respectively, which is somewhat colder than in the last years.

### 2.1.3.2 Zooplankton

In the area north of Iceland, zooplankton biomass tends to be higher in years with strong inflow of Atlantic Water than in years when Atlantic inflow is weak, and lower salinity in the surface layers slows or prevents vertical mixing. A strong inflow of Atlantic water to the north Icelandic area was observed both during November 2002 and February/March 2003. The relatively high zooplankton biomass off the central north coast in spring 2003 is in line with this (Fig. 2.1.3.2.1).

In spring 2003, the zooplankton biomass for the whole Icelandic area was slightly below the long-term average. West of Iceland zooplankton biomass was near average, but slightly below the long term mean south and east of Iceland. The copepod Calanus finmarchicus was generally the dominant zooplankton species in the offshore areas, except in the Arctic East Icelandic current northeast of Iceland, where the arctic copepod $C$. hyperboreus dominated the biomass.

As mentioned above, a continued strong inflow of Atlantic water to the north Icelandic area was observed during last years surveys. On the whole, the zooplankton biomass in Icelandic waters in 2005 was above average. South and west of the country the biomass was near average, but considerably higher off the north and east coasts. As compared to 2004 the zooplankton biomass was higher on most stations.

The NSSP herring which migrated west onto the East-Icelandic shelf south of the cold tongue (approx. $64^{\circ} 50^{\prime}$ ) in late May-early June 2005 was feeding heavily and could easily have stayed on should she have so wished. According to the fishery these herring, however, backed again to the SE in the 2 . week of June, but no systematic search of the area was conducted after that.

### 2.1.4 Hydrography of the waters west of the British Isles

Hydrographic data have been collected during surveys in the spawning season of blue whiting in spring. The mean temperature and salinity from 50 to 600 m of all the stations in deep water (bottom depth>600m) in $2^{\circ}$ latitude times $2^{\circ}$ longitude boxes have been calculated for each survey. The box with limits $52^{\circ}$ to $54^{\circ} \mathrm{N}$ and $16^{\circ}$ to $14^{\circ} \mathrm{W}$ had few gaps, and the time series of mean temperature and salinity for this box is shown in Figure 2.1.4.1. The pattern seen is that after some years with temperatures around $10.1^{\circ} \mathrm{C}$ in the 1980 s, it dropped to a minimum in $1994\left(\sim 9.8^{\circ} \mathrm{C}\right)$. After 1994 an increase in temperature is seen, and in 1998 temperature
reached a local maximum $\left(\sim 10.5^{\circ} \mathrm{C}\right)$ with the three following years a few tenths of a degree colder. 2002 was a warm year with $\sim 10.7^{\circ} \mathrm{C}$, and in 2003 the temperature dropped to $\left(\sim 10.5^{\circ} \mathrm{C}\right) .2004$ was the warmest on record $\left(\sim 10.8^{\circ} \mathrm{C}\right)$, but $2005\left(\sim 10.4^{\circ} \mathrm{C}\right)$ is colder than the three preceding years. This is above the long-term average, but about average for the last 10 years. The increase in temperature coincides with the increase in recruitment of blue whiting. However, it is not know whether there is a causal relationship between hydrographic conditions and recruitment of blue whiting.

### 2.2 Ecosystem impact on the fish stocks

### 2.2.1 Norwegian spring spawning herring

## Feeding and growth

Individual growth of the Norwegian spring spawning herring, as measured by condition or length specific weight after the summer feeding period in the Norwegian Sea, has been characterised by large fluctuations during the 1990’s (Fig 2.2.1.1). During 1991 and 1993 individual condition was good, but from 1994 on the condition of the herring started to decline and by 1997 it reached the lowest level during the 1990's. The level observed in 1997 corresponds with the absolute long-term low level observed during the period 1935 - 1994 (Dr. scient. thesis J.C. Holst 1996, University of Bergen). Following a recovery during 1998 and 1999, the condition of the herring decreased again. During 2001 to 2004 the condition remained at a low level, but slowly increasing.

Since 1995, when the large-scale migration pattern of the herring has been mapped during two annual cruises, May and July-August, the herring have been feeding most heavily in Atlantic water, and the herring condition index obtained after the feeding period in the Norwegian Sea is related to average zooplankton biomass of Atlantic water (Fig. 2.2.1.2). To improve this relationship herring feeding areas should be defined more precisely, because large variations in herring migration routes and in zooplankton distribution have been observed over the years. Extreme changes in migration occurred during the summers 2004 and 2005 when increasing amounts of herring started to feed in the southwestern Norwegian Sea, towards the east coast of Iceland. At the same time we observed that increasing numbers of herring were not overwintering in the fjords of northern Norway, but in the deep waters off the shelf. The herring which are still overwintering inside the fjords had much higher condition than the herring outside, probably due to differences in migration route and feeding conditions between the two groups of overwintering herring. We have used the condition factor of the herring outside the fjord in Fig. 2.2.1.2 (see also 3.9).

A regression of herring condition on the two-months average of the NAO indices showed that the relationship was strongest between herring condition and the NAO during the March-April period (Fig. 2.2.1.3). The prediction for 2005 based on equation (2) is 0.84 , and for 2006, 0.82 , somewhat above and below average, respectively. The condition factor for 2004 was calculated for the fraction of the stock overwintering outside the fjords, and the predictions for 2005 and 2005 are probably valid only for the same part of the stock (see 3.9).

Condition (yr2) $=0.022 *$ NAO yr1 +0.82
$R^{2}=0.51, P=0.004$

## Recruitment

Predictions of the recruitment in fish stocks are essential for future harvesting of the fish stocks. Traditionally, prediction methods have not included effects of climate variability. Multiple linear regression models can be used to incorporate both climate and fish parameters. Especially interesting are the cases where there exists a time lag between the predictor and response variables as this gives the opportunity to make a prediction. A model for the number
of three year old recruits of Norwegian spring spawning herring using the herring 0-group log index and the NCEP skin temperature describes $\sim 80 \%$ of the variation in the recruitment (Figure 2.2.1.4).

The model is:

$$
\operatorname{Re} c_{t}=8.3 \times \text { skin }_{t-3}+16 \times \text { orroup }_{t-3}-44
$$

where Rec is the number (in $10^{9}$ ) of 3 year old recruits of Norwegian spring spawning herring from the WGNPBW 2003 SEASTAR assessment (ICES 2004), skin the NCEP skin (sea surface) temperature in degree C in the Norwegian Sea ( $64-70^{\circ} \mathrm{N}, 6^{\circ} \mathrm{W}-8^{\circ} \mathrm{E}$ ) averaged from January to March 3 years earlier and 0group the 0 -group log index of herring larvae from the survey in the autumn 3 years earlier. The subscripts denote the time lag in years. Further details can be found in Stiansen et al. (2002).

The dominant variable in the model is the 0 -group index, which has a correlation coefficient of 0.84 with the Recruitment (3 years later). When the model was tested on the 0 -group index alone it gave an $R^{2}$ of 0.71 . Still the model explained $9 \%$ more of the variability when adding the skin temperature.

The prognosis shows a steady increase in recruitment for the period 2005-2007, ending at a historic high level in 2007 (Recruits 3 years old: $2005-9.9 * 10^{9}$, $2006-15.8 * 10^{9}$, $2007-$ $26.8 * 10^{9}$ ).

Conclusions:

- Herring condition was lower than average for the time series in 2005.
- There is a weak relationship between zooplankton biomass in May and herring condition in the autumn during the years 1995-2005.
- The March-April NAO index for 2004 and 2005 predicts the herring condition index at 0.84 in the winter 2005 and at 0.82 in the winter 2006.
- Recruitment is predicted to increase during the period from 2005 to 2007.


Figure 2.1.2.1.1. Main surface currents of the Nordic Seas.


Figure 2.1.2.1.2. Hurrell's winter NAO index (Lisbon-Stykkisholmur/Reykjavik), from 1950 to 2004 (blue line), and Osborn's winter NAO index (Gibraltar-Southwest Iceland) from 1995 to 2005 (red line). Black line is 5 years moving averages.


Figure 2.1.2.1.2. Temperature $\left({ }^{\circ} \mathrm{C}\right)$ and salinity observed during July/August, in the core of Atlantic Water beyond the shelf edge in the sections Svinøy - NW, Gimsøy - NW and Sørkapp - W, averaged between 50 and 200 m depth and horizontally over three stations across the core.


Figure 2.1.2.1.3. Time series of area (blue, in km2) and averaged temperature (red/pink) of Atlantic water in the Svinøy section, observed in March/April (triangles) and July/August (dots) 1978-2005.

Ocean Weather Station Mike 2003


Figure 2.1.2.2.1 Distribution of chlorophyll a at 10 m depth during the year at Weather Station Mike in 2003.


Figure 2.1.2.2.2 Year to year variations in the different phases of the development of phytoplankton at Weather Station Mike in the period 1991 to 2003. Diamonds: winter phase; squares: pre-bloom phase; triangles: spring bloom. Continuous lines represent the average for each phase. Broken lines represent one standard deviation for each phase.


Figure 2.1.2.3.1 Zooplankton biomass (dry weight) in the upper 200 m in May. A: Arctic influenced water (salinity $<35$, west of $1.4^{\circ} \mathrm{E}$ ). B: Atlantic water (salinity $>35$ ). B: Norwegian Coastal water (salinity $<35$, west of $1.4^{\circ} \mathrm{E}$ ). Error bars: 95\% confidence limits.


Figure 2.1.2.3.2 Zooplankton biomass in July-August in the eastern Norwegian Sea (0-200 m). Integrated biomass within a fixed geographical region divided by its area.


Figure 2.1.2.4.1. Zooplankton biomass in May, observed and modelled. Model: Biomass (yr2) = 2.23*NAO $\mathbf{y r} 1+10.54$. $\mathrm{R} 2=0.47, \mathrm{P}=\mathbf{0 . 0 2}$. The model predicts a biomass of 10.59 g dry weight $\mathbf{m}-2$ for May 2006.


Figure 2.1.3.1.1. Temperature and Salinity deviations on the Siglunes section north of Icleand, mean for stations 15 and 0 - 200m, 1952 - 2003.


Figure 2.1.3.2.1. Variations in zooplankton biomass (g dry weight $\mathbf{m - 2 , 0} 0-50 \mathrm{~m}$ ) in spring at Siglunes section. The columns show means for 8 stations.


Figure 2.1.4.1. Yearly mean temperature and salinity from 50-600m (crosses) of all stations in a box with bottom depth $>600 \mathrm{~m}$, west of the Porcupine bank bounded by $52^{\circ}$ to $54^{\circ} \mathrm{N}$ and $16^{\circ}$ to $14^{\circ} \mathrm{W}$. Dotted lines are drawn at plusminus one standard deviation of all observations in each box, each year.

Herring condition index


Figure 2.2.1.1 Individual weight to length ratio (herring condition index) for Norwegian spring spawning herring. Data from November and December for herring 30-35 cm body length. Error bars: 95\% confidence limits. In 2004 only herring wintering outside the Ofoten-fjord were used.


Figure 2.2.1.2 Zooplankton biomass (dry weight) in Atlantic water in the Norwegian Sea in May (0-200 m) and herring condition index (individual weight to length ratio, November and December, 30-35 cm). Error bars: 95\% confidence limits. Linear regression: Condition $=0.004$ * biomass $\mathbf{+ 0 . 7 6 6 , ~ R 2 ~}=0.32, \mathrm{P}=0.09$.


Fig. 2.2.1.3 Herring condition index in December, observed and modelled. Model: Condition (yr2) =0.022*NAO $\mathrm{yr} 1+0.82, \mathrm{R} 2=0.51, \mathrm{P}=0.004$. The model predicts herring condition index in December 2005 at 0.844 and in December 2006 at 0.82 .


Figure 2.2.1.4. The figure shows the number of recruits (3 year olds) of Norwegian spring spawning herring (black) and the model fit (red), together with prognoses for 2005-2007 (green).

## 3 Norwegian Spring Spawning Herring

### 3.1 Stock description

The Norwegian spring spawning herring is a highly migratory stock that is distributed throughout large parts of the NE Atlantic during its lifespan. It is a herring type with high number of vertebrae, large size at age, large maximum size, different scale characteristics from other herring stocks and large variation in year-class strength. The herring spawns along the Norwegian west coast in February - March. The larvae drift north and northeast and distribute as 0 -group along the Norwegian coast and in the Barents Sea. The Barents Sea is by far the most important juvenile area for the large yearclasses which forms the basis of the large production potential of the stock. Some yearclasses are in addition distributed into the Norwegian Sea basin as 0-group. Example of this is the 1950 and 2002 yearclasses. Most of the young herring leave the Barents Sea as 3 years old and feed in the northeastern Norwegian Sea for 1-2 years before recruiting to the spawning stock. The adult individuals of the Norwegian spring-spawning herring have a distinct annual migration pattern in the Norwegian Sea. This migration pattern changes over time, at present the herring winters in fjord areas in Northern Norway and off the Vesterålen area from $69^{\circ}$ to $72^{\circ} \mathrm{N}$, spawn on the Norwegian coast (mainly between $62^{\circ}$ and $71^{\circ} \mathrm{N}$ ) and feed in the Norwegian Sea. Since 2003 a more southwestern feeding pattern has been observed with increasing amounts of the older herring feeding in the waters north of the Faroes and east of Iceland. As the feeding season progress the herring has a northerly migration along the polar front zone. During the autumn 2004 NSSH were found in smaller concentrations in catches of Icelandic summer spawning herring off the Icelandic east coast. The total catch of NSSH was estimated at approx. 1000 tonnes. This is a new development which is probably coupled to the more southwestern feeding pattern observed in the recent years and should be followed closely.

The Barents Sea component of the large 2002 yearclass migrated out of the Barents Sea during spring of 2005. It is now found in the waters between west to north of the Vesterålen area and has merged with the Norwegian Sea component of the same yearlclass. The Norwegian Sea component is on average more westerly distributed than the Barents Sea component as a consequence of the size dependent migration pattern.

### 3.2 ICES advice and management applicable to 2004 and 2005

EU, Faroe Islands, Iceland, Norway, and Russia agreed to implement a long-term management plan for Norwegian spring-spawning herring. This plan consists of the following elements (ICES 2002/CRR:255):

1. Every effort shall be made to maintain a level of Spawning Stock Biomass (SSB) greater than the critical level ( $\left.\boldsymbol{B}_{\text {lim }}\right)$ of $2500000 t$.
2. For the year 2001 and subsequent years, the Parties agreed to restrict their fishing on the basis of a TAC consistent with a fishing mortality rate of less than 0.125 for appropriate age groups as defined by ICES, unless future scientific advice requires modification of this fishing mortality rate.
3. Should the SSB fall below a reference point of $5000000 t\left(\boldsymbol{B}_{p a}\right)$, the fishing mortality rate, referred under paragraph 2, shall be adapted in the light of scientific estimates of the conditions to ensure a safe and rapid recovery of the SSB to a level in excess of $5000000 t$. The basis for such an adaptation should be at least a linear reduction in the fishing mortality rate from 0.125 at $\boldsymbol{B}_{p a}\left(5000000\right.$ t) to $0.05 \boldsymbol{B}_{\text {lim }}(2500000$ t).
4. The Parties shall, as appropriate, review and revise these management measures and strategies on the basis of any new advice provided by ICES.

ICES considers that the objectives of this agreement are consistent with the precautionary approach.

In 2003 ACFM stated that "Based on the most recent estimate of SSB and fishing mortality ICES classifies the stock as being inside safe biological limits. The stock is harvested around $\mathbf{F}_{\mathrm{pa}}=0.15$. The recruitment of the very strong 1992 year class led to an increase in SSB in 1997 to approximately 8 million t , but SSB has since declined to just over 5 million t in 2002. The incoming year classes 1998 and 1999 are estimated to be relatively strong." Further "ICES advises that this fishery should be managed according to the agreed management plan with a fishing mortality of no more than $\mathrm{F}=0.125$, corresponding to landings in 2004 of less than 825000 t."

At the meeting on Fisheries Consultation on the management of Norwegian spring-spawning herring (Atlanto-Scandian) herring stock in Reykjavik, Iceland in October 2003, the coastal states (European Union, Faroe Islands, Iceland, Norway, and Russia) did not reach any agreement regarding the allocation of the quota.

At a following meeting on Fisheries Consultation on the management of Norwegian springspawning herring stock in Copenhagen, Denmark, in mid February 2004, the parties were unable to reach any agreement on quota allocations. However, there seemed to be an unwritten understanding between the parties to accept the TAC proposed by ACFM to limit the total catches to 825000 t in 2004.

In 2004 ACFM stated that " Based on the most recent estimates of SSB and fishing mortality, ICES classifies the stock as having full reproduction capacity and harvested sustainably. The recruitment of the very strong 1992 year class led to an increase in SSB in 1997 to approximately 8 million t . Thereafter, SSB declined to just below 5 million t in 2001 and increased again to 7 million t in
2004. The year classes 1998 and 1999 are estimated to be relatively strong." Further " The management plan implies catches of 890000 t in 2005 which is expected to lead to spawning stock of 6.3 million tonnes in 2005."

For 2005 there was no agreement between the Coastal States regarding the allocation of the quota. The Nowegians rose their quota of $14 \%$ and following them up so did the Icelanders and the Faroese. The sum of the total revised national quotas for 2005amounts to about 1 million tonnes.

### 3.3 Description and development of the fisheries

The distribution of the fisheries of Norwegian spring-spawning herring by all countries in 2004 by ICES rectangles are shown in Figure 3.3.1 (total whole year) and in Figure 3.3.2 (by quarter).

Due to limitations by some countries to enter the EEZs of other countries in 2004 the fisheries do not necessarily depict the distribution of herring in the Norwegian Sea and the preferred fishing pattern of the fleets give free access to any zone. However, in general the development of the international fishery shown by these figures follows the known migration pattern for Norwegian spring-spawning herring. The migration pattern, together with environmental factors, was mapped in 2004 and 2005 during the ICES PGNAPES (Planning Group on Surveys on Pelagic Fish in the Norwegian Sea) investigations (ICES 2004/D:10 and ICES 2005/D:09).

### 3.3.1 Denmark

5. The Danish fishery of Norwegian spring spawning herring is carried out by purse seiners ( $18,000 \mathrm{t}$.) and trawlers ( $5,000 \mathrm{t}$.). Most of the landings were landed in Denmark ( $11,000 \mathrm{t}$.) and the other landings were landed at the Faroes $(10,000 \mathrm{t}$.) and the remaining part in Iceland ( $2,000 \mathrm{t}$.). No landings were landed in Norway do to lack of agreements between EU and Norway. In 2004 the first fishing
period started in the southern part of Division IIa in March (catch 3,600 t) and continued in May - June were app. 17,300 t was caught. Finally the landings from the second part of the year were minor.

### 3.3.2 Germany

The German fishery for Norwegian spring spawning herring started in Division IIa north of $66^{\circ} \mathrm{N}$ in late April and continued in May-June in IIa and further north in IIb with the highest catches taken in end of June. Less catches were taken in third quarter in the Svalbard area.

### 3.3.3 Faroe Islands

The Faroese fishery for Norwegian spring-spawning herring (7 combined purse seiners/trawlers) started in late May 2004 in International waters (ICES Division IIa) close to the Faroese fishing border. The fishery continued in summer gradually moving further northeast in the international area and into the Svalbard area (IIa and IIb). In August the fishery was almost exclusively taken in the Svalbard area (IIb). More than $60 \%$ of the catches were taken with pelagic trawls and the rest with purse seines.

### 3.3.4 Iceland

The Icelandic catch quota for Norwegian spring-spawning herring was set at 128205 tonnes in 2004. The Icelandic fishery in 2004 began early May in the western part of the international waters in the Norwegian Sea between the Jan Mayen and Faroese zones. Later in the month the fishery continued there as well as in north-eastern part of the international and the eastern part of the Icelandic zone In June the fishery moved into the southern Spitsbergen zone and continued there until the last week of August when it moved into the international waters again. In September the fishery remained in the international waters close to the Spitsbergen zone. When the Icelandic fishery on summer spawners began in autumn some Norwegian spring spawning herring was mixed with the Icelandic summer spawning herring stock off East Iceland. It was estimated that close to 900 t of NSSH were caught in autumn at East Iceland during the fishery of the Icelandic summer spawning herring. In May and June about 4500 t of NSSH were caught in Icelandic EEZ. The bulk of the catch was caught in May and June ( 56 thousands tonnes) and in July and August ( 32 thousands tonnes) and a few tonnes were caught in September. The total catch was 102787 tonnes of which 68 thousands tonnes were taken in midwater trawl and 34 thousands tonnes in purse-seine.

About 25 purse-seiners/trawlers participated in the herring fishery. The length range of the vessels was 47-79 meters with a mean length of 63 meters. The engine power range of the fleet was 441-5520 kW (599-7500 HP) with a mean of 2751 kW ( 3738 HP ).

### 3.3.5 Ireland

The Irish catches in 2004 were insignificant.

### 3.3.6 Netherlands

The Dutch fleet fishing for pelagic species in European waters consists of 14 freezer trawlers and one pair trawler. In addition, a number of flag vessels are operating from the Netherlands. Target species of this fleet are: herring, blue whiting, mackerel, horse mackerel and argentines Some of these trawler also operate in west African waters during part of the year fishing for sardinella and horse mackerel. The fishery for Norwegian spring spawning is conducted using large pelagic trawls. In 2004, 8 vessels have participated in the fishery and reported catches from 17 trips. Most of the herring catches originate from a directed fishery in the second half of the year in the months June-October in ICES Sub-area II (International area in the Norwegian Sea).

### 3.3.7 Norway

The Norwegian fishery is carried out by many size categories of vessels. Of the total national quota of 470.250 t , approximately $50 \%$ is allocated to purse seiners, $10 \%$ to trawlers and $40 \%$ to smaller coastal purse seiners. Like in previous years the by far larger part of the Norwegian fishery takes place in northern Norwegian coastal waters (Vestfjord area) where the herring winters from September until mid January. In 2004 about 115000 tonnes were caught in the wintering areas in Northern Norway in January. 80000 t were taken in the spawning area on the Norwegian coast in February-March. Only negligible quantities ( 452 tonnes) were caught in the areas south of 62 N in 2004. Much of this herring probably belongs to local fjordic herring stocks but is registered as NSSH in the statistical records. The remaining part of the Norwegian quota (approximately 277000 t ) was taken in the period September-December on the herring migrating to, and wintering in, the wintering areas in northern Norway. The total Norwegian catch in 2003 was 476624 tonnes.

### 3.3.8 Russia

In 2004 the Russian fishery started within the shelf region of the Norwegian EEZ, near Trena Bank (approximately $66^{\circ} \mathrm{N}$ ) in the beginning of February and Sclinna Bank (approximately $65^{\circ} \mathrm{N}$ ) and Buagrunnen Bank (approximately $63^{\circ} \mathrm{N}$ ) in the end of this month. In March the fishing was in progress in the same regions. In February and March the catch was 17608 t . In June the commercial vessels conducted fishing in the northern part of the international area in the Norwegian Sea and the zone of Spitsbergen. In June the catch was 11227 t . In July-August vessels caught herring in the international area in the Norwegian Sea in the Polar Front region and the zone of Spitsbergen. In the middle of August Russian vessels followed the southward migrating fish and continued their fishery in the Norwegian EEZ. In September the fishery of the herring was prolonged in the EEZ of Norway. The herring migrated southwestwards, along the depths of the continental slope. In July-September the catch was 70792 t . In November Russian fishery finished in the Norwegian EEZ to north from Lofoten area where was caught 13939 t. The Russian fishery is carried out by many types of vessels, mainly trawls. The entire Russian catch was utilized for human consumption.

### 3.3.9 Sweden

The Swedish fishery for Norwegian spring-spawning herring was performed by 15 boats using purse seines ( $60 \%$ of landings) and pelagic trawls ( $40 \%$ ). Catches were taken during June and July around the border between IIa and IIb and landed on the Faroes (60\%), in Denmark and Sweden.

### 3.3.10 UK (Scotland)

The Scottish fishery for NSSH takes place at the end of the first quarter, once the mackerel season has finished and also at the end of the third quarter, once the North Sea herring fishery has been completed. Around one half of the pelagic fleet participates in this fishery using either single or pair trawl gear.

### 3.4 Bycatches in the fishery

No information on bycatches in the fishery was supplied to the working group.

### 3.5 Fishery dependent data

### 3.5.1 Sampling intensity

For the direct fisheries of the herring in 2004 samples were provided by Denmark, Iceland, Norway and Russia. The sampled catch accounted for $78,6 \%$ of the total catch. Better sampling $(85,8 \%)$ of catch in ICES area IIa, worse sampling ( $6,8 \%$ ) in area IIb and no samples were taken of the catches in areas IVa and Vb .

The Working Group noted that not all nations participating in the international fishery for Norwegian spring spawning herring in 2004 had carried out an adequate sampling of their fishery. The allocation of catches for which no samples were taken and the final catch-at age and weight- at age by ICES areas is given in Table 3.5.1.1. In general Norwegian age distribution and weight were used for un-sampled fishery in the Norwegian Sea in quarter 2 for IIa and 1-4 IVa area ICES, Russian age distribution and weight keys for quarter 3-4 for IIa area ICES for un-sampled fishery in 1 and 3 quarter for for area IIb and Danish data for 2 quarter for area ICES IIb.

### 3.5.2 Landings

Like in earlier years the fishing pattern in 2004 followed the clockwise migration pattern of the herring (Figure 3.3.3.1). Two main changes in the fishing pattern could be distinguished from 2003: There was no fishing in the Jan Mayen zone in 2004 since no coastal state agreement had been reached allowing non Norwegian vessels to fish in the area. There was also a tendency for more catches being taken in the southwestern areas, in line with the observed increase in concentrations of herring feeding in this area. The herring feeding in these western areas are known to be composed of the largest fish in the stock due to the length dependent migration behaviour of herring. The tonnage taken in the two areas compose a small fraction of the total catch and the changes in fishery pattern have probably not changed the selection pattern appreciably.

The total annual catches of Norwegian spring-spawning herring for the period 1972-2003 (2003 preliminary) are presented in Table 3.5.2.1 (by country). In 2004 the catch provided as catch by rectangle represented approximately 799236 tonnes, 7606 tonnes higher than the total catch used in the WG (Figure 3.3.1). Iceland revised their catch upwards 2036 tonnes at the WG, and the catch from UK (Scotland) was 7345 tonnes higher in the catch-by-rectangle format than in the reported forms.

The Working Group noted that in this fishery an unaccounted mortality caused by fishing operations and underreporting probably exists. In general, it was not possible to assess the magnitude of these extra removals from the stock, and taking into account the large catches taken in recent years, the relative importance of such additional mortality is probably low. Therefore, no extra amount to account for these factors has been added in 1994 and later
years. In previous years, when the stock and the quotas were much smaller, an estimated amount of fish was added to the catches.

The Norwegian water content in catches was set at $2 \%$ as of 1 February 2004 by the Norwegian authorities. This came after an international agreement on application of a standard water content in the herring catches. No extra tonnage was consequently added to the official Norwegian catch in 2004 (see last years report for an elaborate description of the water content history).

### 3.5.3 Discards

In 2004 the part of a young Norwegian spring spawning herring (2002 year-class) was allocated in the eastern part of Norwegian Sea, therefore there is a possibility for by-catch small fish in this area. It is possible that discard of small fish takes place. However the Working Group has no accessible data to estimate possible discards of the herring.

### 3.5.4 Age and length composition of catches

Age compositions for direct fishery of the herring in the Norwegian Sea were provided by Denmark, Iceland, Norway and Russia, and the sampled catch accounted for $89 \%$ of the total catch. Estimates of catch in numbers for unsampled catches were raised according to the knowledge of how, where, and when the catches were taken. The age compositions in the directed fisheries are given in Table 3.5.1.1 and Table 3.5.4.1. These data were used in the stock assessment. The 1999-1998 year classes were the most numerous in the catches, followed by the 1992-1991 year classes. To calculate the total international catch-at-age, and to document how it was done, the program SALLOC was used (ICES 1998/ACFM:18).

Data on the combined length composition of the 2004 commercial catch by quarter of the year from the directed fisheries in the Norwegian Sea were provided by Iceland, Norway and Russia. Length composition of the herring varied from 13 to 41 cm , with $97 \%$ of fish ranging from 28-37 cm. The mean length in this fishery was 32.7 cm (Table 3.5.4.2).

### 3.5.5 Weight at age

Weight at age in stock for 2005 (January $1^{\text {st }}$ ) (Table 3.5.5.1) was taken from the weights at age in the Norwegian samples made in the wintering areas in December 2004. Prognosis of weight at age in stock for 2006 was estimated by adjusting the weights at age for 2005 with the condition prognosis given for December 2005 (sec. 2.2.1). Correspondingly the prognosis of weight at age in stock for 2007 was estimated by adjusting the weights at age for 2005 with the condition prognosis given for December 2006 (sec. 2.2.1).

The weight in catches in 2004 was taken from the total international weight-at-age (Table 3.5.5.2), which were produced using the computer programme SALLOC, standard ICES software.

### 3.5.6 Length at age

Not used.

### 3.5.7 Maturity at age

Maturation of the 2002 yearclass: The 2002 yearclass is composed of three components with separate juvenile areas: 1) The Barents Sea 2) The Norwegian Sea 3) Norwegian fjords. The bulk of the yearclass ( $\sim 97-98 \%$ ) is made up of the two oceanic components. Of these two the Barents Sea component is the largest, composing about $75-80 \%$ of the yearclass. Due to the small fraction made up of the fjordic component only the Norwegian Sea and Barents Sea
components were directly included in the calculation of the maturity ogive of the yearclass. The Barents Sea component will start maturing from 4 years of age while the Norwegian Sea component started as 3 years old. The maturity ogive for the yearclass in 2005 was calculated as follows: It was assumed that $0 \%$ of the Barents Sea component would mature in 2005. Of the Norwegian Sea component all fishes in the December 2004 survey in stage 3 were supposed to mature (Observations: Stage 2: 67\%, Stage 3:33\%). The maturation of the 2002 yearclass in 2005 thus becomes: $0.75 * 0+0.25 * 0.33=0.085$. This number was raised to 0.1 , taking into account some fish from the coastal component, which matures at a low age (Table 3.5.7.1).

For other ages and yearclasses the maturity ogive were kept like last year for 2005.

### 3.5.8 Natural Mortality

No changes were done to the applied natural mortalities, 0.9 for 0 to 2 years old, 0.15 for 3 years and older.

### 3.6 Fisheries independent data

### 3.6.1 Survey abundance indices

Due to the change of timing of the NPBWWG from spring to autumn 2005, both the May survey 2004 and in 2005 are included in the present description of surveys, as well as the recent summer surveys in the Norwegian Sea and the Barents Sea.

### 3.6.1.1 Spawning grounds

In 2005 a Norwegian acoustic survey was again undertaken to estimate the abundance of herring in the spawning areas in February-March. During the years 2001-2004 the obtained estimates were either of a bad quality or the survey was not carried out (Table 3.6.1.1.1). The age groups 5-15+ are used in the assessment.

### 3.6.1.2 Wintering areas

Norwegian acoustic surveys have been carried out in the wintering areas in November/December (1992-present) and in January (1991-1999).

The wintering area of the herring is now split in two areas. This was first observed during the December survey in 2002 when concentrations of herring were observed by the RV Johan Hjort in oceanic waters off Vesterålen and Troms. In December 2004 a survey covering both the fjordic and oceanic wintering area was carried out by the RV G.O.Sars (only fjords) and R/V Johan Hjort. The combined result of the survey (Table 3.6.1.2.1) covers the known wintering area of the mature part of the stock. There was a very distinct difference in age structure between the two areas with most of the 1998 and 1999 yearclasses wintering in the oceanic areas and hardly any of the older yearclasses wintering in these areas. The age groups $4-15+$ are used in the assessment.

The results of the survey in the wintering area in January is found in Table 3.6.1.2.2. Although the survey series has ended the data are still used in the assessment. The age groups $5-15+$ are used in the assessment.

### 3.6.1.3 Feeding areas

The feeding areas in the Norwegian Sea were surveyed acoustically by the international ecosystem survey (PGNAPES, former PGPFN) during the period late April to early June in

2004 and 2005 (ICES 2004/D:07 and ICES 2005/D:09). A complete herring data set is found in the PGNAPES reports from the years 1995-2005 at, -WW.imrong estimate is given in Table 3.6.1.3.1. The age groups 3-15+ are used in the assessment.

### 3.6.1.4 Nursey areas

The nursery areas of the Norwegian spring-spawning herring are Norwegian fjords and coastal areas, and in the Barents Sea. Since 1988, when the 1983-year class spawned for the first time, the latter area has increased in importance as a nursery area for the herring.

Results from the Norwegian and Russian acoustic survey in the Barents Sea in May-June 1991-2005 are given in Table 3.6.1.4.1. No surveys were carried out in the years 2003-2004. The age groups 1 and 2 are used in the assessment.

In 2001 the Working Group decided to include data on immature herring obtained during the Russian-Norwegian survey in August-October in estimating the younger year classes in the Barents Sea. The results from these surveys are given in Table 3.6.1.4.2. The age groups 1 and 2 are used in the assessment.

The results from the 0 -group herring survey in Norwegian Fjords and Coastal areas are given in Table 3.6.1.4.3, however, the data are not used in the present assessment of herring.

The results from the joint Norwegian-Russian 0-group survey in the Barents Sea are given in Table 3.6.1.4.4. The log index is used in the assessment.

### 3.6.1.5 Herring larval survey

The Norwegian herring larval survey in 2005 was carried out during March-April. The survey started at Fugløya $\left(70^{\circ} \mathrm{N}\right)$ and continued along the Norwegian shelf south to Stad $\left(62^{\circ} \mathrm{N}\right)$. No herring larvae were found in this area. The mean size of the larvae was relatively low, 11.5 mm , which may be due to the early survey period in 2005 . The oldest larvae were found in the south indicating an earlier start of the spawning in this area. The larvae had a more northern distribution compared to the historic distribution of this survey. The total number of larvae was estimated to be $73.9 * 10^{12}$, the highest estimate so far in this series (Table 3.6.1.5.1). The "Index 1 " is used in the assessment.

### 3.6.2 Tagging data

With the exception of 1999, 2001 and 2005, tagging has been carried out annually since 1975 by Norway. The tagging experiments in 2004 were carried out in March along the Norwegian coast from $62^{\circ} \mathrm{N}-67 \mathrm{~N}$ where a total of 27045 herring were tagged. During the tagging process, the length of each tagged herring is measured. For each purse seine catch that is used for tagging, a sample of 100 fish is taken to determine the age distribution within each length group. The age composition of tagged herring in this batch is then estimated from the age distribution in the sample.

Recovery of tags from supervised detector plants has continued, as well as recovery from the standard magnets in the production line of fish processing plants and from individuals. For stock assessment purposes, tags are only used from supervised detector plants where detector efficiency has been tested, and where it is known that the detectors have been working as intended. Three factories filled these criteria in 2004, and a total of 54321 million herring were screened in these factories. Magnet efficiency was close to $100 \%$ in 2004. All tagged herring recovered were sent to the Institute of Marine Research, Bergen, where they were measured, weighed and aged. In 200484 tags from herring that were four years or more when tagged, were recovered from the factories (Table 3.2.4.1).

### 3.7 Stock Assessment

Stock assessment was carried out with SeaStar and ISVPA.

### 3.7.1 Catch curve analyses

Not carried out.

### 3.7.2 Data Exploration with assessment models

Data exploration was carried out with SeaStar and ISVPA.

### 3.7.2.1 Data Exploration: Sea Star

In SeaStar assessments the catchabilities of the acoustic surveys are assumed to have no age or abundance dependence. The reason for this choice is that the mechanism for the recruiting of herring to the surveys is that of migration, not of growth, where the latter process could be assumed to be more regular and modellable. Instead, for each survey there is set a minimum age for inclusion of each year class. However, one needs to be cautious, since the different year classes may have different recruitment ages to the surveys. An important example is the failure of the 1998 year class to recruit fully to the December survey in 2002, for which reason the WG meeting 2003 considered this data point an outlier. In the coming years special attention should be paid also to the large 2002 year class, where a substantial part grew up in the Norwegian Sea instead of in the Barents Sea.

However, the recruitment to the Norwegian Sea survey by the 1998 and 1999 year classes (see Figure 3.7.2.1.5) may indicate an age dependent catchability for that survey, which should be considered modelled in future assessments.

The input data used are:
Catch data Updated from Table 3.5.1.1
Acoustic surveys
Tag data
Table 3.6.1.1-5

Larval data
0 -group data

Table 3.6.2
Table 3.6.1.5
Table 3.6.1.4.4

The acoustic surveys used are (the numbering is used elsewhere in the text of this section):

| 1 | Spawning grounds along the Norwegian coast | Minimum age: 5 |
| :--- | :--- | :--- |
| 2 | Wintering area in Vestfjorden in November-December | Minimum age: 4 |
| 3 | Wintering area in Vestfjorden in January | Minimum age: 5 |
| 4 | Young herring in the Barents Sea in May | Ages 1 and 2 |
| 5 | Feeding areas in the Norwegian Sea in May | Minimum age: 3 |
| 6 | As part of the joint IMR-PINRO capelin survey in September | Ages 1 and 2 |

It is assumed that the distribution of the main tuning series of older fish follows a gamma distribution with a common CV, which is estimated, and that the distribution of the acoustic data in the Barents Sea follows a log normal distribution. The tag return data are assumed to follow a Poisson distribution, which is commonly used for rare events, the larval data are assumed to follow a gamma distribution with an estimated CV and the zero group data are assumed to follow a normal distribution with an estimated standard deviation.

It has been experienced that when the fish get old the scales get difficult to read and more scales get discarded. This introduces a bias in the age distribution. An attempt has been made to correct for this bias on an experimental basis in previous WG meetings. This was not done during the present meeting, however.

Previously it has been observed in this WG that as the 1983 and 1985 year classes grew older than about 13 years the age readers tended to transfer fish from the 1983 to the 1985 year class. This problem is part of the general age reading problem for older fish, but is corrected for separately in the SeaStar software. The data used in ISVPA have also this correction.

As has been the case during previous assessments of Norwegian spring spawning herring with SeaStar, a number of exploratory runs were initially performed to see the effect of various options and settings. These runs were:

| Label | Explanation |
| :---: | :---: |
| Default | With respect to the 2004 assessment the 2002 year class was included in the tuning |
| The other runs are deviations from Default: |  |
| NoTags | Tag data were left out |
| EstimateM | The natural mortality both for adult herring and young herring in the Barents Sea were estimated |
| IncrMPlus | The natural mortality in the plus group was set to 0.5 |
| TrendLarvae | An attempt to correct for a time trend in the larval data |
| NoLarv | The larval data were left out |
| Log | The surveys in the Norwegian Sea and along the Norwegian coast are assumed distributed as normal on log-scale |
| 2002Bar | The measurement of the 2002 year class in the Norwegian Sea is left out, th estimate of this year class is based on Barents Sea data- this run has identical settings to the final SeaStar run in 2004 |
| S2And5NoLNoT | No tags, no larvae, only surveys 2 and 5 |

The estimated parameters, catchabilities and stock biomass in 2005 are shown in Table 3.7.2.1.1 for the exploratory runs. The catchabilities reflect how well each survey conforms to the modelled stock.

The perceived spawning stock is somewhat increased when the tagging data are removed and decreased when the larval data are removed. If $M$ is estimated there is a decrease in the spawning stock of about one million tonnes. The M-value for adult herring is estimated at 0.103 as opposed to 0.115 in 2004 and 0.128 in 2003, i.e. there has thus been a slightly decreasing trend in the estimated M-value. The M-value for young herring in the Barents Sea is estimated at 1.117 , as opposed to 0.188 in 2004 and 0.507 in 2003, i.e a substantial fluctuation from year to year. These fluctuations are probably connected to removing the 1998 and 1999 year classes from the Barents Sea likelihood (stage 2 of the estimation) in 2004 and the 2002 year class in 2005.

Increasing M in the plus group caused only a slight decrease in present spawning stock size, but reduced the spawning stock in the collapsed phase in the 1970s to below 0.1 million tonnes. The WG chose this run as the SeaStar assessment in 2005. Figure 3.7.2.1.1 shows the spawning stock biomass, figure 3.7.2.1.2-7 the survey fits, figure 3.7.2.1.8 the fit to the larval data. Figure 3.7.2.1.9 shows the quantile-quantile plot for the surveys and figure 3.7.2.1.10 the quantile-quantile plot for tags. Figure 3.7.2.1.11 shows the quantile-quantile plot for the surveys when a log-normal distribution is assumed.

The quantile-quantile plots for the tag return data shows a skewed distribution. There are more points with low CDF than expected, i.e. The model expects more tag returns than it gets, so the tag data will tend to move the assessment towards higher tag return rates, i.e. towards a lower stock. This is consistent with a increased stock when the tag data are removed.

The larval data seem to be a reasonably good proxy for the spawning stock, except that since 1997 there seems to be a regime of higher levels. Introducing a linear trend in the data does not improve the fit. These data should be reviewed before the next meeting, looking into possible explanations for the discrepancy between the larval indices before 1997 and since 1997 in relation to modelled spawning stock. Since the WG could not point to any biological reasons for this, the WG decided to keep this survey in the tuning at also at the present assessment. However, these data should be looked closer into with the aim of relating the discrepancy to survey coverage in relation to fluctuating spawning areas, timing of survey with respect to the time of spawning in each year or other reasons.

A gamma distribution seem to fit the acoustic survey data better than a log-normal distribution, but the latter does not seem entirely inappropriate.

Table 3.7.1.2.2 shows a comparison of number by age in 2004 as perceived by the assessment made in 2004 and the present assessment. Of the strong year classes, the 2002 year class has been substantially revised upwards, the perception of the 1998 and 1999 year classes is virtually unchanged and the 1991 and 1992 year classes have been modestly revised downwards.

It may be biologically justifiable that the M -value of the plus group is higher than the M -value for younger fish. The selected value of 0.5 is, however, admittedly speculative. There is strong evidence that the smallest spawning stock biomass in the 1970s could not be as high as 0.5 million tonnes, which is what we get with an M of 0.15 on the plus group. An alternative could be to base the calculation of the numbers in the plus group on the catch and an assumed F-value equal to the F-value of the oldest true age group. This is, however, a very strong assumption.

As the increased M -value of the plus group did not affect the present assessment appreciably, the WG felt that the M-value of 0.5 reflected the stock history better, for which reason this run was chosen.

Figure 3.7.2.1.12 shows the retrospective plot. There has been substantial a downward revisions of the perceived stock history from 1999 until 2004. However, the present
assessment increases the perceived stock history. Figure 3.7.2.1.13 shows the estimated catchabilities in the retrospective runs. The catchabilities for the surveys in the Norwegian Sea and along the Norwegian coast show a smooth development, reflecting that the model around 2000 assumed a higher stock relative to the surveys than before or after.

### 3.7.2.2 Data Exploration: ISVPA

For NSS herring exploratory runs by means of ISVPA the same version of the model and the same settings as in last year stock assessment were used (the catch-controlled version of the ISVPA with constraint of unbiased model approximation of logarithmic catch-at-age, two selection patterns were fitted: first - for 1950-1985, and second - for 1986-2003), except the following two aspects:

Profiles of components of ISVPA loss function for different sources of information are represented on Figure3.7.2.2.1. (It is necessary to mention that the last year report there was a a misspecification in picture 6 of the analogous figure - the median was minimized for survey 6 , not the sum of squared errors).

Survey in feeding area in May (N5 on figures) same way as in previously, was used for tuning in form of age proportions, weighted by stock abundance. Weighting by abundance is used in order to make stress on abundant years as probably giving survey information of better quality.

The only marked difference between last year and this year profiles is that now the minimum for survey 1 is shifted towards higher values of fishing mortality.

For the sake of comparison with the SeaStar assessment, where several points in survey data are "masked" (excluded, because by some reasons some of them are treated to be in contradiction to the others), an additional run was made using survey data with the SeaStar "mask" applied. Profiles of the ISVPA loss function components obtained using "masked" survey data are presented on fig.3.7.2.2.2. The only marked influence of this was shift of the minimum for surveys 5 , but it did not influence the position of the minimum for the total loss function, and, accordingly, the overall solution (compare curves 7 on Figures3.7.2.2.1 and 3.7.2.2.2).

An attempt to detect outliers in surveys data was also undertaken on the basis of the SPALY ISVPA run results. The so called "X-84" rule by P.Huber was applied. According to this rule the points with residuals higher than 5.2 absolute median deviations are to be excluded (Vasilyev, 2004). Using this rule the following points were detected as outliers and excluded:

| Source of information | \% of outliers with respect to all points <br> in the data source | age (and years) of <br> outliers |
| :--- | :---: | :--- |
| Survey 1 (spawning, Feb) | 1.4 | $9(1995)$ |
| Survey 2 (Wintering, Nov) | 0 |  |
| Survey 3 (Wintering, Jan) | 3.2 | $9(1991)$ <br> $11(1993)$ |
| Survey 4 (Young, Barents <br> Sea, June) | 0 |  |
| Survey 5 (Feeding, May) | 5.6 | $4(2005)$ |
|  |  | $9(1996)$ |
|  |  | $10(1996)$ |
|  |  | $14(2000)$ |
| Survey 6 (Young, Barents | 0 | $16(2001,2005)$ |
| Sea, Sep) |  |  |

X-84-rule is a very "soft" distribution-free criterion intended to detect apparent outliers; its "softness" helps to avoid false detection and not to worry about the organization of an iterative winsorization-like procedure of correction of outliers, which for more "rigorous" criteria may require many iterations and may move the whole assessment into false direction.

Figure 3.7.2.2.3 illustrates the influence of the outliers' exclusion on the respective components of the ISVPA loss function. As it can be seen, survey 1 is now less pooling the solution to improper high F; for survey 3 almost no changes are detected (changes are not visible in graphic representation, but very small exist if to look at digits); for survey 5 position of the minimum is not changed, but level of errors is lower and shape of loss function is somewhat more smooth. This may indicate that in general for surveys 1 and 5 exclusion of outliers leads to proper direction, but the median and the absolute median deviation (surveys 3 ) in this case are already sufficiently robust measures of scale.

If to return to experiments with outliers, detected and coming from Seastrar and from ISVPA, it ought to be reminded that the concept of outliers is intrinsically model-dependent. Convinced classic likelyhoodists, who believe that for any kind of data there exists a "true" distribution, may say that they may detect outliers using only postulated proper model of distribution, correspondingly are dependent on this model; but if even they are working with residuals, they also depend on an observation model which influences on these residuals. But if to believe that anything is possible in a sample of limited size and that dealing with a very limited sample it is hardly possible to detect the best distribution model, to may be preferable to use distribution-free robust statistics and methods for detection of outliers.

Thus, after the above mentioned, being short of time to make other experiments, which were also planned in spirit of merging of the ISVPA and The SeaStar model properties and features, for this year assessment by means of ISVPA it was decided to use the same model settings as in 2004. Results of the ISVPA model (in the same settings as last year) application to herring data in their "default" form are given on Figure 3.7.2.2.4.

Comparison of the results with the result of previous assessment are given on Figure 5. They are very much inline. Current assessment shows further spawning stock biomass stock increase in 2004 and slight decrease in 2005.

Results of retrospective runs are given on Figure. 3.7.2.2.6.
Residuals for all sources of data are collected on Figure 3.7.2.2.7.
Since profiles of components of ISVPA loss function for the SPALY run revealed three sources of information with more distinct minima then the others, it was decided to make an additional run using only these three sources: catch-at-age, survey 2 and survey 5 . The results of the run are compared to the results of SPALY run on Figure 3.7.2.2.8.

Since in this run the sources of information with weaker signals, which were in contradiction to more clear signals from catch-at-age and surveys 2 and 5 were excluded, this run was chosen as final ISVPA run.

Residuals for ISVPA run with 3 sources of information are given on Fig. 3.7.2.2.9. Fig. 3.7.2.2.10 represents bootstrap results.

### 3.7.3 Comparison of results of different assessments

The two models used, ISVPA (catch controlled version) and Seastar give somewhat different perception about the state of the stock today with the spawningstock in 2005 estimated to be 11 million tonnes in ISVPA. Historically the models agree reasonably well as expected as both models are VPA type models where age disaggregated catches take over in the past.

In both the models surveys that have been discontinued were used in the tuning. Investigation of the ISVPA SSE profiles indicated that a survey that was discontinued in 1999 (wintering area in January) was having much effect on F in 2004 towards lower F and therefore bigger spawningstock. Tuning only with the December survey on the spawning grounds and the May survey in the Norwegian sea lead to considerable downwards revision of the spawning stock in both models.

An important difference between the two models lies in treatment of the small yearclasses where Seastar does not use those yearclasses in tuning but ISVPA uses all yearclasses. The stock is charaterized by very high contrast in strength of yearclasses which in connection with sampling problems in the surveys leads to very high contrast in numbers in different agegroups in the surveys. Catchabilities are estimated approximately as the mean of log-ratios between survey and number in stock. The observation model used for the large observations may not apply to low observations which may depend on a few age readings. As a consequence, the low observations may lead to wrongly estimated catchabilities which when applied to the large year classes may lead to serious bias in the assessment.

In Seastar estimated catchability is closer to 1 as Seastar does only rely on the large year classes. This is probably the cause for lower estimated catchabilities and lower estimate of recent large year classes (Figure 3.7.3.1). Looking at other catchability related issues like the age after which catchability is independent of age can also matter.

In ISVPA the 1998 and 1999 year classes are stronger than the 1991 and 1992 year classes at the same ages. In SeaStar the 1991 and 1992 year classes are stronger than the 1998 and 1992 year classes, which is in accordance with the survey observations.

The contrast in the age disaggregated survey abundance indices is a problem that has to be taken into account when doing assessment for this stock. Ignoring the small year classes as done in Seastar is not an ideal solution but much better than doing nothing. A quick fix to the problem is to use high detection level in the surveys but the detection level is an indication of how large the survey indices are when sampling errors are no longer the dominant errors and lognormal (or gamma) errors can be assumed. A small check in a VPA model run on this stock showed that changing this detection level by a factor of 20 could change the spawning stock in 2005 by more than 3 million tonnes by estimating higher catchability in the surveys.

As future solution tuning with the total biomass in the survey as lognormal and proportion in each age group as multinomial would also be worth exploring. Also it might be questioned if catchability in the December acoustic survey should be independent of age, but allowing the catchabilities to be estimated for each age group seems to predict decreased catchability with age. Experiments in mackerel very clearly shows that sampling with trawl in an area fished with purse seine gives lower mean lengths in the trawl catches catch than the less selective purse seine catching specific schools. Estimates of catchability are though uncertain for this stock as the survey series are relatively short for a stock with so low fishing and total mortality as estimated for the Norwegian spring spawning herring. Therefore the acoustic surveys need to be scruitinized properly with regard to whether catchability in them could be substantially different from 1 and this belief should possibly be put in the models as a prior.

For more than ten years, tagging data has been used in the assessment of the Norwegian spring spawning herring. SeaStar can handle these data, and therefore when doing the assessment of the stock the tagging data were included as a tuning series. In this assessment, the larvae data are also included. The larvae data is an SSB index, which has also been used for many years in the assessment. Also, the WG considered that the relation between the 1991/1992 and 1998/1999 year classes in the surveys is better reflected in SeaStar than in ISVPA. After an overall evaluation, also including that this assessment is an update from last years the Sea Star assessment was chosen as the final.

### 3.7.4 Final Assessment

The settings for the final run is described in sec 3.2.1. In order to make the historical biomass during the 1970 'ies more in accordance with the observed low stock level at that time the F for the + group was increased from 0.15 to 0.5 in this years assessment.

Table 3.7.1.2.2 shows a comparison of number by age in 2004 as perceived by the assessment made in 2004 and the present assessment. Of the strong year classes, the 2002 year class has been substantially revised upwards, the perception of the 1998 and 1999 year classes is virtually unchanged and the 1991 and 1992 year classes have been modestly revised downwards.

Figure 3.7.2.1.12 shows the retrospective plot. There has been substantial a downward revisions of the perceived stock history from 1999 until 2004. However, the present assessment increases the perceived stock history. Figure 3.7.2.1.13 shows the estimated catchabilities in the retrospective runs. The catchabilities for the surveys in the Norwegian Sea and along the Norwegian coast show a smooth development, reflecting that the model around 2000 assumed a higher stock relative to the surveys than before or after.

The results of the final assessments as stock numbers are given in Table 3.7.4.1. Table 3.7.4.2 shows the stock summary table.

### 3.8 Recruitment estimates

Recruitment indices are listed in the historic summary table (Tables 3.7.4.2). The mean of the estimated recruits in 1950-2004 is 95 billion as 0 year old fish.

The perception of the strong 2002 year class is based on the measurement in the 2005 survey in the Norwegian Sea. The perception of the 2003 year class is based on the 0 -group index and the Barents Sea surveys and the perception of the 2004 year class is based mainly on the 0 group index, which is about the level of the 0-group index for the 1991 year class, and only the 1983 year class ha a higher 0-group index.

### 3.9 Forecast

Besides the short term forecast two optional runs were made due to special circumstances related to the recruitment of the strong 2002 yearclass. The rationale for this and procedure applied is described in $\sec$ 3.9.1.

### 3.9.1 Short term forecast

Table 3.9.1.1 shows the input data to the short-term prediction. Weight at age used in the forecast is described in sec. 3.5.5. The exploitation pattern was chosen as the populationweighted mean of F by age in the 10 last years of catch data.

The catches calculated for 2006 are sensitive to the selection of 4 year olds because of the strong 2002 year class, and one may get different results depending on how long averaging period is selected. The 2002 year class is special in that a part of it grew up in the Norwegian Sea, in contrast to all other recent strong year classes that grew up in the Barents Sea. The exploitation pattern assumed for 2006 may thus be dubious, as the 2002 year class will recruit earlier than other corresponding year classes. Since the 2002 year class may be more strongly represented in the fishery in 2006 than assumed by the exploitation pattern, the short time forecast will give unrealistic low landings (given the state of the stock) in 2006 with the agreed $\mathrm{F}=0.125$ ( 640000 tonnes). This is a problem, which may typically appear the same year as a strong year class start recruiting heavily. For the 2002 yearclass the problem is even more pertinent because of its early maturation (proportion mature is 0.6 in 2006 as 4 years
old). The working group therefore decided to present two additional optional forecasts where alternative selection patterns for the 2002 year class were used in order to arrive at predicted landings in better accordance with the current state of the stock.

Short term: Selection pattern is the population weighted mean of F by age for the last 10 years

Option 1: Selection on 4 year olds as on 5 year olds
Option 2: $\quad$ Regression of catch to historic Fbar
Table 3.9.1.2 shows the short-term prediction tables for the short term, option 1 and option 2 forecast runs.

The upper table shows the standard short term prediction table for the given exploitation pattern. $\mathrm{F}_{\text {bar }}$ equal to 0.125 gives a catch in 2006 of 640000 tonnes. It should be noted that with this exploitation pattern there will be virtually no fishing on the 2002 year class in 2006. Also, $\mathrm{F}_{\text {bar }}$ is based on ages 5-14, and the F-value applied for the 2002 year class in 2006 will not affect $\mathrm{F}_{\text {bar }}$.

The middle table shows the short term prediction table corresponding to assuming the selection on 4 year old herring is the same as the selection of 5 year old herring. This is the same as assuming the 2002 year class is recruited earlier to the fishery than normally. In this case an $\mathrm{F}_{\text {bar }}$ of 0.125 corresponds to a catch in 2006 of about 740.000 tonnes.

The WG has also interpreted the harvesting control rule of $\mathrm{F}=0.125$ differently from applying an $\mathrm{F}_{\text {bar }}$ (population weighted) based on ages 5-14 directly. Figure 3.9.1.1 shows a regression of the ratio of total catch to $\mathrm{F}_{\mathrm{bar}}$ on data from 1990 and later. The regression line was used to calculate the catch for given values of " $\mathrm{F}_{\text {bar }}$ " and partitioning the catch on age groups accorded to the chosen selection pattern. In this case an " $\mathrm{F}_{\text {bar }}$ " of 0.125 corresponds to a catch in 2006 of about 800.000 thousand tonnes.

The present harvesting control rule does not devise a higher F-value in cases where a proven strong year class is about to recruit to the fishery. The consequences of such an adaptive strategy should be tested by simulation, which was impossible for the WG to carry out because of time constraints.

### 3.10 Biological reference points

|  | ICES considers that: | ICES proposed <br> that: |
| :--- | :--- | :--- |
| Precautionary <br> reference points | Approach | $\mathbf{B}_{\text {lim }}$ is 2.5 million t |
|  | $\mathbf{F}_{\text {lim }}$ is not considered relevant for <br> this stock | $\mathbf{B}_{\text {pa }}$ be set at 5.0 <br> million t |
| $\mathbf{F}_{\mathrm{pa}}$ be set at $\mathrm{F}=$ <br> 0.15 |  |  |

## Target Reference Points

Management has defined $\mathrm{F}_{\mathrm{y}}$ at 0.125

## Technical basis:

$$
\begin{array}{|l|l|}
\hline \mathbf{B}_{\mathrm{lim}}: \text { MBAL } & \mathbf{B}_{\mathrm{pa}}:=\mathbf{B}_{\mathrm{lim}} * \exp \left(0.4^{*} 1.645\right) \text { (ICES Study Group 1998) } \\
\hline
\end{array}
$$

| $\mathbf{F}_{\text {lim }}:-$ | $\mathbf{F}_{\mathrm{pa}}:$ ICES Study Group 1998 |
| :--- | :--- |

### 3.11 Management considerations

There has been no international agreement on quota allocations the two last years. This has led to an escalation in the F exerted on the stock, with the fisheries in 2005 probably ending close to 1 million tonnes, close to 100.000 tonnes more than the TAC recommended under the longterm management plan. It is of great importance for a high long-term yield in this stock that the intentions of the management plan are not violated.

### 3.12 Quality of the data and the assessment

For the first time the assessment was carried out after the summer, which means that the May survey result was included in the assessment just after it was carried out and not the year after. It also means that two new May surveys were included in the data set in the present assessment. It is regarded a great advantage with significant effect to the quality of the assessment that the May survey is included shortly after it was carried out and not 11 months after.

The herring is at present moving both its wintering areas and partly its summer feeding area. These dynamics could lead to new factors affecting the survey results. For the wintering area the present coverage is several orders lower in magnitude than it used to be when the entire adult stock wintered in the fjords. In the fjords the extent of the wintering area was fairly well known before the survey started and the surveys could be optimised with regard to the herring distribution. With the oceanic wintering area this is no more the case. Based on the fisheries a rough understanding of the geographic distribution of the herring could still be obtained before the November-December wintering survey, thus helping somewhat in planning the survey transects, but in no way compensating fully.

In can be seen from Figure 3.7.2.12 that there is a strong retrospective pattern in the assessment, in overestimating the spawning stock biomass. This is a negative factor in the assessment indicating an unstable assessment. This tendency has though stopped this year.

In sum there are both positive and negative factors affecting the quality of the present assessment. Probably the inclusion of two May surveys and that the assessment is done shortly after the survey results in an assessment of better quality than last year.

### 3.13 Recommendations

The present harvesting control rule does not devise a higher F-value in cases where a proven strong year class is about to recruit to the fishery. The consequences of such an adaptive strategy should be tested by simulation, which was impossible for the WG to carry out because of time constraints.

Table 3.5.1.1
Summary of Sampling by Country - Norwegian spring spawning herring

AREA : Vb

| Country | Sampled <br> Catch | Official <br> Catch | No. of <br> samples | measured | No. | aged |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |

AREA : IVa

| Country | Sampled <br> Catch | Official <br> Catch |
| :--- | ---: | ---: |
| UK (Scot) | 0.00 | 405.00 |
| Norway | 0.00 | 452.00 |
| Ireland | 0.00 | 11.00 |
| Total IVa | 0.00 | 868.00 |
|  |  | 868.00 |
| Sum of Offical Catches : | 0.00 |  |
| Unallocated Catch : |  | 868.00 |

AREA : IIb

| Country | Sampled Catch | Official Catch | No. of samples | No. <br> measured | No. aged | $\begin{gathered} \text { SOP } \\ \% \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sweden | 0.00 | 5195.00 | 0 | 0 | 0 | 0.00 |
| Russia | 0.00 | 2310.00 | 0 | 0 | 0 | 0.00 |
| Iceland | 4340.00 | 37335.00 | 1 | 47 | 50 | 860.24 |
| Germany | 0.00 | 4152.00 | 0 | 0 | 0 | 0.00 |
| Faroes | 0.00 | 18165.00 | 0 | 0 | 0 | 0.00 |
| Denmark | 502.00 | 3947.00 | 4 | 477 | 477 | 785.40 |
| Total IIb | 4842.00 | 71104.00 | 5 | 524 | 527 | 852.48 |
| Sum of Offical Catches :Unallocated Catch : |  | 71104.00 |  |  |  |  |
|  |  | 0.00 |  |  |  |  |
| Working Group Catch : |  | 71104.00 |  |  |  |  |


| Country | Sampled <br> Catch | Official <br> Catch |
| :--- | ---: | ---: |
| UK(Scot) | 0.00 | 1464.00 |
| Sweden | 0.00 | 2791.00 |
| Russia | 102339.00 | 113566.00 |
| Norway | 476624.00 | 476624.00 |
| Netherlands | 0.00 | 17369.00 |
| Iceland | 32722.00 | 63416.00 |
| Germany | 0.00 | 658.00 |
| Faroes | 0.00 | 24556.00 |
| Denmark | 5740.00 | 19164.00 |
| Total IIa | 617425.00 | 719608.00 |

$$
\begin{array}{lr}
\text { Sum of Offical Catches : } & 719608.00 \\
\text { Unallocated Catch : } & 0.00 \\
\text { Working Group Catch : } & 719608.00
\end{array}
$$

PERIOD : 1


> Official Catch 17608.00 193394.00 7056.00 218058.00

Sum of Offical Catches : Unallocated Catch :
Working Group Catch :

PERIOD : 2

| Country | Sampled Catch | Official Catch | No. of samples | No. <br> measured | No. aged | $\begin{aligned} & \text { SOP } \\ & \% \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK (Scot) | 0.00 | 1464.00 | 0 | 0 | 0 | 0.00 |
| Sweden | 0.00 | 5043.00 | 0 | 0 | 0 | 0.00 |
| Russia | 0.00 | 11227.00 | 0 | 0 | 0 | 0.00 |
| Norway | 4106.00 | 4175.00 | 5 | 150 | 285 | 100.01 |
| Iceland | 37062.00 | 55344.00 | 16 | 717 | 848 | 271.85 |
| Germany | 0.00 | 3934.00 | 0 | 0 | 0 | 0.00 |
| Faroes | 0.00 | 18411.00 | 0 | 0 | 0 | 0.00 |
| Denmark | 6242.00 | 15173.00 | 6 | 724 | 724 | 243.10 |
| Period Total | 47410.00 | 114771.00 | 27 | 1591 | 1857 | 253.19 |
| Sum of Offical | ches : | 114771.00 |  |  |  |  |
| Unallocated Cat |  | 0.00 |  |  |  |  |
| Working Group C |  | 114771.00 |  |  |  |  |

PERIOD : 3

| Country | Sampled Catch | Official Catch |
| :---: | :---: | :---: |
| Sweden | 0.00 | 2943.00 |
| Russia | 70792.00 | 73102.00 |
| Norway | 94303.00 | 94382.00 |
| Netherlands | 0.00 | 10533.00 |
| Iceland | 0.00 | 40823.00 |
| Germany | 0.00 | 876.00 |
| Faroes | 0.00 | 24310.00 |
| Denmark | 0.00 | 238.00 |
| Period Total | 165095.00 | 247207.00 |
| Sum of Offical Catches :Unallocated Catch : |  | 247207.00 |
|  |  | 0.00 |
| Working Group Catch : |  | 247207.00 |

PERIOD : 4

| Country | Sampled Catch | Official Catch |
| :---: | :---: | :---: |
| UK (Scot) | 0.00 | 405.00 |
| Russia | 13939.00 | 13939.00 |
| Norway | 185006.00 | 185125.00 |
| Netherlands | 0.00 | 6836.00 |
| Ireland | 0.00 | 11.00 |
| Iceland | 0.00 | 4584.00 |
| Faroes | 0.00 | 50.00 |
| Denmark | 0.00 | 644.00 |
| Period Total | 198945.00 | 211594.00 |
| Sum of Offical Ca | ches : | 211594.00 |
| Unallocated Catch |  | 0.00 |
| Working Group Cat | h | 211594.00 |


| No. of <br> samples | No. <br> measured | No. <br> aged | SOP <br> $\%$ |
| :---: | :---: | :---: | ---: |
| 0 | 0 | 0 | 0.00 |
| 5 | 1982 | 318 | 100.25 |
| 162 | 9093 | 2683 | 100.00 |
| 0 | 0 | 0 | 0.00 |
| 0 | 0 | 0 | 0.00 |
| 0 | 0 | 0 | 0.00 |
| 0 | 0 | 0 | 0.00 |
| 0 | 0 | 0 | 0.00 |
| 167 | 11075 | 3001 | 100.02 |
|  |  |  |  |

Total over all Areas and Periods

| Country | Sampled <br> Catch | Official <br> Catch | No. of <br> samples | measured | No. | aged |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |

## Table 3.5.1.1 continued

DETAILS OF DATA FILLING-IN

| Filling-in | for record : ( 5) | Norway | 1 | IVa |
| :---: | :---: | :---: | :---: | :---: |
| Using Only |  |  |  |  |
| >> ( 1) | Norway | 1 IIa |  |  |
| Filling-in | for record : ( 6) | Norway | 2 | IVa |
| Using Only |  |  |  |  |
| >> ( 1) | Norway | 1 IIa |  |  |
| Filling-in | for record : ( 7) | Norway | 3 | IVa |
| Using Only |  |  |  |  |
| >> ( 1) | Norway | 1 IIa |  |  |
| Filling-in | for record : ( 8) | Norway | 4 | IVa |
| Using Only |  |  |  |  |
| >> ( 1) | Norway | 1 IIa |  |  |
| Filling-in | for record : ( 10) | Russia | 2 | IIa |
| Using Only |  |  |  |  |
| >> ( 11) | Russia | 3 IIa |  |  |
| Filling-in | for record : ( 13) | Russia | 3 | IIb |
| Using Only |  |  |  |  |
| >> ( 11) | Russia | 3 IIa |  |  |
| Filling-in | for record : ( 14) | Denmark | 1 | IIa |
| Using Only |  |  |  |  |
| >> ( 1) | Norway | 1 IIa |  |  |
| Filling-in | for record : ( 16) | Denmark | 3 | IIa |
| Using Only |  |  |  |  |
| >> ( 3) | Norway | 3 IIa |  |  |
| Filling-in | for record : ( 17) | Denmark | 4 | IIa |
| Using Only |  |  |  |  |
| >> ( 12) | Russia | 4 IIa |  |  |
| Filling-in | for record : ( 18) | Denmark | 1 | IIb |


| Using Only | Russia | 1 IIa |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Filling-in | for record : ( 21) | Iceland | 3 | IIa |
| $\begin{array}{cc} \text { Using Only } \\ \gg & (20) \end{array}$ | Iceland | 2 IIa |  |  |
| Filling-in | for record : ( 22) | Iceland | 4 | IIa |
| $\begin{array}{cc} \text { Using Only } \\ \gg & (20) \end{array}$ | Iceland | 2 IIa |  |  |
| Filling-in | for record : ( 24) | Iceland | 3 | IIb |
| $\begin{gathered} \text { Using Only } \\ \gg \\ (11) \end{gathered}$ | Russia | 3 IIa |  |  |
| Filling-in | for record : ( 25) | Sweden | 2 | IIa |
| $\begin{array}{cc} \text { Using Only } \\ \gg & \left(\begin{array}{rl} 2 \end{array}\right) \end{array}$ | Norway | 2 IIa |  |  |
| Filling-in | for record : ( 26) | Sweden | 3 | IIa |
| $\begin{array}{cc} \text { Using Only } \\ \gg & (11) \end{array}$ | Russia | 3 IIa |  |  |
| Filling-in | for record : ( 27) | Sweden | 2 | IIb |
| $\begin{array}{cc} \text { Using Only } \\ \gg & (19) \end{array}$ | Denmark | 2 IIb |  |  |
| Filling-in | for record : ( 28) | Sweden | 3 | IIb |
| $\begin{array}{cc} \text { Using Only } \\ \gg & (11) \end{array}$ | Russia | 3 IIa |  |  |
| Filling-in | for record : ( 29) | Germany | 2 | IIa |
| $\begin{array}{cc} \text { Using Only } \\ \gg & (2) \end{array}$ | Norway | 2 IIa |  |  |
| Filling-in | for record : ( 30) | Germany | 2 | IIb |
| $\begin{array}{cc} \text { Using Only } \\ \gg & (19) \end{array}$ | Denmark | 2 IIb |  |  |
| Filling-in | for record : ( 31) | Germany | 3 | IIb |



## Table 3.5.1.1 continued

Catch Numbers at Age by Area

| Ages | Vb | IVa | IIb | IIa | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.00 | 0.00 | 0.00 | 125.00 | 125.00 |
| 1 | 0.00 | 0.00 | 0.00 | 1814.40 | 1814.40 |
| 2 | 1.92 | 3.67 | 2189.10 | 40940.26 | 43134.95 |
| 3 | 4.61 | 6.70 | 5421.85 | 25453.09 | 30886.26 |
| 4 | 35.04 | 48.72 | 32805.59 | 110003.00 | 142892.36 |
| 5 | 165.04 | 275.96 | 325479.09 | 551713.31 | 877633.44 |
| 6 | 172.87 | 709.95 | 425810.75 | 772300.88 | 1198994.63 |
| 7 | 18.10 | 153.46 | 66510.80 | 128939.09 | 195621.44 |
| 8 | 5.85 | 230.59 | 28530.12 | 133842.73 | 162609.33 |
| 9 | 0.45 | 23.46 | 5046.79 | 28113.38 | 33184.07 |
| 10 | 2.38 | 57.43 | 14532.73 | 52615.46 | 67208.01 |
| 11 | 5.53 | 179.78 | 24823.88 | 159865.53 | 184874.72 |
| 12 | 9.09 | 665.48 | 32472.07 | 378488.78 | 411635.41 |
| 13 | 4.19 | 357.28 | 12217.76 | 196579.25 | 209158.47 |
| 14 | 1.49 | 24.96 | 2594.90 | 24975.90 | 27597.25 |
| 15 | 0.32 | 31.98 | 548.78 | 10584.00 | 11165.08 |

Mean Weight at Age by Area (Kg)

| Ages | Vb | IVa |
| :---: | ---: | ---: |
| 0 | 0.0000 | 0.0000 |
| 1 | 0.0000 | 0.0000 |
| 2 | 0.1084 | 0.0733 |
| 3 | 0.1764 | 0.1304 |
| 4 | 0.1853 | 0.1711 |
| 5 | 0.2127 | 0.2132 |
| 6 | 0.2308 | 0.2630 |
| 7 | 0.2641 | 0.2964 |
| 8 | 0.3004 | 0.3292 |
| 9 | 0.3082 | 0.3320 |
| 10 | 0.3189 | 0.3562 |
| 11 | 0.3273 | 0.3648 |
| 12 | 0.3324 | 0.3726 |
| 13 | 0.3447 | 0.3825 |
| 14 | 0.3629 | 0.4056 |
| 15 | 0.4303 | 0.4171 |

Ib
0.0000
0.0000
0.1360
0.1983
0.2124
0.2264
0.2394
0.2761
0.3116
0.3510
0.3253
0.3459
0.3418
0.3659
0.3677
0.4620

| IIa | Total |
| :---: | :---: |
| 0.0224 | 0.0224 |
| 0.0655 | 0.0655 |
| 0.1416 | 0.1413 |
| 0.1801 | 0.1833 |
| 0.2111 | 0.2114 |
| 0.2423 | 0.2364 |
| 0.2762 | 0.2631 |
| 0.3078 | 0.2970 |
| 0.3496 | 0.3429 |
| 0.3600 | 0.3586 |
| 0.3719 | 0.3618 |
| 0.3887 | 0.3829 |
| 0.3931 | 0.3890 |
| 0.4011 | 0.3990 |
| 0.4251 | 0.4197 |
| 0.4383 | 0.4394 |

Table 3.5.2.1
Total catch of Norwegian spring-spawning herring (tonnes) since 1972. Data provided by Working Group members.

| Year | Norway | USSR/ <br> Russia | Denmark | Faroes | Iceland | Ireland | Netherlands | Greenl and | UK | Germany | France | Poland | Sweden | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 13,161 | - | - | - | - | - | - | - | - | - | - | - |  | 13,161 |
| 1973 | 7,017 | - | - | - | - | - | - | - | - | - | - | - | - | 7,017 |
| 1974 | 7,619 | - | - | - | - | - | - | - | - | - | - | - |  | 7,619 |
| 1975 | 13,713 | - | - | - | - | - | - | - | - | - | - | - | - | 13,713 |
| 1976 | 10,436 | - | - | - | - | - | - | - | - | - | - | - | - | 10,436 |
| 1977 | 22,706 | - | - | - | - | - | - | - | - | - | - | - | - | 22,706 |
| 1978 | 19,824 | - | - | - | - | - | - | - | - | - | - | - | - | 19,824 |
| 1979 | 12,864 | - | - | - | - | - | - | - | - | - | - | - | - | 12,864 |
| 1980 | 18,577 | - | - | - | - | - | - | - | - | - | - | - | - | 18,577 |
| 1981 | 13,736 | - | - | - | - | - | - | - | - | - | - | - | - | 13,736 |
| 1982 | 16,655 | - | - | - | - | - | - | - | - | - | - | - | - | 16,655 |
| 1983 | 23,054 | - | - | - | - | - | - | - | - | - | - | - | - | 23,054 |
| 1984 | 53,532 | - | - | - | - | - | - | - | - | - | - | - | - | 53,532 |
| 1985 | 167,272 | 2,600 | - | - | - | - | - | - | - | - | - | - | - | 169,872 |
| 1986 | 199,256 | 26,000 | - | - | - | - | - | - | - | - | - | - | - | 225,256 |
| 1987 | 108,417 | 18,889 | - | - | - | - | - | - | - | - | - | - | - | 127,306 |
| 1988 | 115,076 | 20,225 | - | - | - | - | - | - | - | - | - | - | - | 135,301 |
| 1989 | 88,707 | 15,123 | - | - | - | - | - | - | - | - | - | - | - | 103,830 |
| 1990 | 74,604 | 11,807 | - | - | - | - | - | - | - | - | - | - | - | 86,411 |
| 1991 | 73,683 | 11,000 | - | - | - | - | - | - | - | - | - | - | - | 84,683 |
| 1992 | 91,111 | 13,337 | - | - | - | - | - | - | - | - | - | - | - | 104,448 |
| 1993 | 199,771 | 32,645 | - | - | - | - | - | - | - | - | - | - | - | 232,457 |
| 1994 | 380,771 | 74,400 | - | 2,911 | 21,146 | - | - | - | - | - | - | - | - | 479,228 |
| 1995 | 529,838 | 101,987 | 30,577 | 57,084 | 174,109 | - | 7,969 | 2,500 | 881 | 556 | - | - | - | 905,501 |
| 1996 | 699,161 | 119,290 | 60,681 | 52,788 | 164,957 | 19,541 | 19,664 | - | 46,131 | 11,978 | - | - | 22,424 | 1,220,283 |
| 1997 | 860,963 | 168,900 | 44,292 | 59,987 | 220,154 | 11,179 | 8,694 | - | 25,149 | 6,190 | 1,500 | - | 19,499 | 1,426,507 |
| 1998 | 743,925 | 124,049 | 35,519 | 68,136 | 197,789 | 2,437 | 12,827 | - | 15,97119, | 7,003 | 605 | - | 14,863 | 1,223,131 |
| 1999 | 740,640 | 157,328 | 37,010 | 55,527 | 203,381 | 2,412 | 5,871 | - | 207 | - | - | - | 14,057 | 1,235,433 |
| 2000 | 713,500 | 163,261 | 34,968 | 68,625 | 186,035 | 8,939 | - | - | 14,096 | 3,298 | - | - | 14,749 | 1,207,201 |
| 2001 | 495,036 | 109,054 | 24,038 | 34,170 | 77,693 | 6,070 | 6,439 | - | 12,230 | 1,588 | - | - | 9,818 | 766,136 |
| 2002 | 487,233 | 113,763 | 18,998 | 32,302 | 127,197 | 1,699 | 9,392 | - | 3,482 | 3,017 | - | 1,226 | 9,486 | 807,795 |
| 2003 | 438,140 | 122,846 | 14,144 | 27,943 | 117,910 | 1,400 | 8,678 | - | 9,214 | 3,371 | - | - | 6,431 | 750,077 |
| $2004{ }^{1}$ | 477,076 | 115,876 | 23,111 | 42,771 | 102,787 | 11 | 17,369 | - | 1,869 | 4,810 |  | - | 7,986 | 793,666 |

${ }^{1}$ Preliminary, as provided by Working Group members.

## Table 3.5.4.1. Catch data for Norwegian spring spawning herring, billion individuals.

Year
1950
1951
1952
1953
1954
1955
1956
1957
1958
1959
1960
1961
1962
1963
1964
1965
1966
1967
1968
1969
1970
1971
1972
1973
1974
1975
1976
1977
1978
1979
1980
1981
1982
1983
1984
1985
1986
1987

| 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 113 | . 000 | 0.600 | 0.276 | 0.185 | 0.186 | 0.547 | 0.629 | 0.080 | 0.089 | 0.110 | 0.087 | 0.195 | 0.368 | 0.066 | 0.107 | 0. |
| 1.636 | 7.608 | 0.400 | 0.007 | 0.384 | 0.172 | 0.164 | 0.516 | 0.602 | 0.077 | 0.083 | 0.103 | 0.108 | 0.254 | 0.348 | 0.047 | 5 |
| 13.720 | 9.150 | 1.233 | 0.039 | 0.061 | 0.602 | 0.136 | 0.205 | 0.380 | 0.378 | 0.079 | 0.086 | 0.108 | 0.107 | 0.187 | 0.256 | 0.308 |
| 5.697 | 055 | 0.581 | 0.740 | 0.047 | 0.101 | 0.356 | 0.082 | 0.111 | 0.314 | 0.395 | 0.062 | 0.091 | 0.094 | 0.099 | . 216 | 0.515 |
| 10.680 | 7.071 | 0.855 | 0.266 | 1.436 | 0.1 | 0.236 | 0.490 | 0.128 | 0.200 | 0.440 | 0.461 | 0.088 | 0.101 | 0.133 | 0.127 | 0. |
| 5.176 | 2.871 | 0.510 | 0.093 | 0.276 | 2.045 | 0.114 | 0.190 | 0.275 | 0.085 | 0.193 | 0.296 | 0.203 | 0.059 | 0.085 | 104 | 0.477 |
| 5.364 | 2.024 | 0.627 | 0.117 | 0.252 | . 314 | 555 | 0.110 | 0.204 | 0.264 | 0.131 | 0.198 | 0.27 | 0.163 | 0.063 | . 089 | 0.476 |
| 5.002 | 3.291 | 0.220 | 0.023 | 0.373 | 0.154 | 0.229 | 1.985 | 0.072 | 0.127 | 0.183 | 0.088 | 0.121 | 0.149 | 0.132 | 0.034 | 48 |
| 9.667 | 798 | . 666 | 0.018 | 0.018 | 0.111 | 0.089 | 0.194 | 0.973 | 0.071 | 0.123 | 0.201 | 0.099 | 0.077 | 0.071 | . 069 | 0.186 |
| 17.900 | 0.199 | 0.326 | 0.015 | 0.027 | 0.026 | 0.147 | 0.115 | 0.241 | 1.104 | 0.089 | 0.124 | 0.198 | 0.089 | 0.077 | 0.085 | 0.151 |
| 12.880 | 13.580 | 0.393 | 0.122 | 0.018 | 0.028 | 0.024 | 0.096 | 0.073 | 0.204 | 1.163 | 0.085 | 0.130 | 0.154 | 0.057 | 0.047 | 22 |
| 6.208 | 16.080 | 2.885 | 0.031 | 0.008 | 0.004 | . 15 | 0.019 | 0.062 | 0.049 | 0.136 | 0.728 | 0.05 | 0.045 | 0.063 | 0.022 | 38 |
| 3.693 | 4.081 | 1.041 | 1.844 | . 008 | 0. | 07 | 0.020 | 0.012 | 0.059 | 0.053 | . 117 | 0.814 | 0.044 | 0.055 | . 66 | 7 |
| 4.807 | 2.119 | 2.045 | 0.760 | 0.836 | 0.005 | 0.002 | 0.004 | 0.018 | 0.009 | 0.108 | 0.093 | 0.17 | 0.924 | 0.080 | . 060 | 25 |
| 3.613 | 2.728 | 0.220 | 0.115 | 0.399 | 2. | 0.014 | 0.002 | 0.003 | 0.025 | 0.029 | . | 0. | 0.153 | 0.773 | 46 | 0.291 |
| 2.303 | 3.781 | 2.854 | 0.090 | 0.256 | 0.571 | 2.200 | 0.020 | 0.015 | 0.007 | 0.019 | 0.040 | 0.101 | 0.108 | 0.139 | 0.704 | 0.179 |
| 3.927 | 0.663 | 1.678 | 2.049 | 0.027 | . 46 | 06 | 2.885 | 0.038 | 0.014 | 0.017 | 0.026 | 0.01 | 0.06 | 0.072 | 0.097 | 60 |
| 0.427 | 9.877 | 0.070 | 1.392 | 3.254 | 0.027 | 0.421 | 1.132 | 1.721 | 0.009 | 0.006 | 0.004 | 0.008 | 0.009 | 0.018 | 0.014 | . 090 |
| 1.784 | 0.437 | 0.388 | 0.099 | 1.881 | 1 | 0.014 | 0.094 | 0.134 | 0.345 | 0.002 | 0.001 | 0.000 | 0.003 | 0.003 | . 02 | 0.015 |
| 0.561 | 0.507 | 0.142 | 0.188 | 0.000 | 0.009 | 0.005 | 0.000 | 0.012 | 0.034 | 0.036 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 |
| 0.119 | 0.529 | 0.033 | 0.006 | 0.019 | 0.000 | 03 | 0.003 | 0.001 | 0.013 | 0.026 | 0.028 | 0.000 | 0.000 | 0.000 | . 00 | 02 |
| 0.031 | 0.043 | 0.085 | 0.002 | 0.001 | 0.0 | 0.000 | 0.001 | 0.001 | 0.000 | 0.004 | 0.007 | 0.005 | 0.000 | 0.000 | . 00 | 00 |
| 0.347 | 0.041 | 0.020 | 0.035 | 0.003 | 0.004 | 0.002 | 0.000 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 00 |
| 9 | 0.004 | 0.002 | 0. | 0.025 | 0. | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0 | 0.000 | 0.000 | 0 | 0 |
| 0.066 | 0.008 | 0.004 | 0.000 | 0.000 | 0.025 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0. | 0.0 | 0.002 | 0.003 | 0 | 0.000 | 0.031 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | . 00 | 00 |
| 0.020 | 0.002 | 0.001 | 0.023 | 0.005 | 0.000 | 0.000 | 0.013 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0 |
| 0. | 0.006 | 3 | 0.022 | 0.024 | 0. | 0.000 | 0.000 | 0.011 | 0.000 | 0.000 | 0. | 0.000 | 0.000 | 0.000 | . 00 | 00 |
| 0.020 | 0.002 | 0.001 | 0.003 | 0.012 | 0.020 | 0.000 | 0.000 | 0.000 | 0.005 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | . 0.00 |
| 0.033 | 0.004 | 0.002 | 0.006 | 0.002 | 0.007 | 0.011 | 0.000 | 0.000 | 0.000 | 0.003 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.007 | 0.000 | 0.000 | 0.006 | 0.006 | 0.002 | 0.008 | 0.016 | 0.000 | 0.000 | 0.000 | 0.003 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.008 | 0.001 | 0.012 | 0.004 | 0.005 | 0.00 | . 00 | 0.005 | 0.008 | 0.000 | 0.000 | 0.000 | 0.00 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.023 | 0.001 | 0.000 | 0.014 | 0.008 | 0.005 | 0.006 | 0.002 | 0.005 | 0.006 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | . 000 |
| 0.127 | 0.005 | 0.002 | 0.003 | 0.021 | 0.010 | 0.006 | 0.007 | 0.001 | 0.005 | 0.007 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.034 | 0.002 | 0.002 | 0.004 | 0.005 | 0.062 | 0.018 | 0.013 | 0.016 | 0.007 | 0.016 | 0.006 | 0.000 | 0.000 | 0.000 | 0.002 | . 000 |
| 0.029 | 0.013 | 0.207 | 0.022 | 0.016 | 0.017 | 0.130 | 0.059 | 0.055 | 0.063 | 0.010 | 0.031 | 0.050 | 0.000 | 0.000 | 0.000 | 0.003 |
| 0.014 | 0.001 | 0.003 | 0.540 | 0.018 | 0.01 | 0.016 | 0.105 | 0.075 | 0.042 | 0.077 | 0.019 | 0.066 | 0.080 | 0.000 | 0.000 | 2 |
| 14 | 0.006 | 0.036 | 0.020 | 0.501 | 0.019 | 0.004 | 0.007 | 0.028 | 0.012 | 0.010 | 0.005 | 0.008 | 0.007 | 0.007 |  |  |

## Table 3.5.4.1. Catch data for Norwegian spring spawning herring, billion individuals. (con't)

| 1988 | 0.015 | 0.003 | 0.009 | 0.063 | 0.025 | 0.550 | 0.009 | 0.004 | 0.006 | 0.015 | 0.009 | 0.003 | 0.003 | 0.003 | 0.002 | 0.000 | 0.000 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1989 | 0.007 | 0.002 | 0.025 | 0.003 | 0.004 | 0.006 | 0.324 | 0.003 | 0.000 | 0.000 | 0.003 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1990 | 0.001 | 0.000 | 0.016 | 0.019 | 0.003 | 0.012 | 0.011 | 0.226 | 0.001 | 0.002 | 0.002 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1991 | 0.000 | 0.003 | 0.003 | 0.008 | 0.003 | 0.001 | 0.015 | 0.009 | 0.219 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1992 | 0.002 | 0.000 | 0.001 | 0.013 | 0.033 | 0.005 | 0.001 | 0.012 | 0.006 | 0.226 | 0.002 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.000 |
| 1993 | 0.007 | 0.000 | 0.007 | 0.028 | 0.107 | 0.087 | 0.009 | 0.004 | 0.030 | 0.019 | 0.410 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1994 | 0.000 | 0.000 | 0.008 | 0.033 | 0.110 | 0.364 | 0.165 | 0.016 | 0.008 | 0.037 | 0.036 | 0.645 | 0.003 | 0.000 | 0.000 | 0.002 | 0.000 |
| 1995 | 0.000 | 0.000 | 0.001 | 0.058 | 0.346 | 0.623 | 0.638 | 0.231 | 0.016 | 0.016 | 0.070 | 0.084 | 0.912 | 0.004 | 0.000 | 0.000 | 0.000 |
| 1996 | 0.000 | 0.000 | 0.030 | 0.034 | 0.714 | 1.571 | 0.941 | 0.406 | 0.103 | 0.006 | 0.007 | 0.066 | 0.018 | 0.837 | 0.000 | 0.000 | 0.000 |
| 1997 | 0.000 | 0.000 | 0.022 | 0.130 | 0.271 | 1.796 | 1.994 | 0.761 | 0.326 | 0.061 | 0.020 | 0.032 | 0.091 | 0.019 | 0.370 | 0.000 | 0.000 |
| 1998 | 0.000 | 0.000 | 0.083 | 0.070 | 0.242 | 0.368 | 1.760 | 1.264 | 0.381 | 0.130 | 0.043 | 0.025 | 0.003 | 0.113 | 0.006 | 0.109 | 0.000 |
| 1999 | 0.000 | 0.000 | 0.005 | 0.138 | 0.036 | 0.135 | 0.429 | 1.605 | 1.164 | 0.291 | 0.106 | 0.015 | 0.040 | 0.007 | 0.089 | 0.000 | 0.064 |
| 2000 | 0.000 | 0.000 | 0.014 | 0.084 | 0.560 | 0.035 | 0.111 | 0.404 | 1.299 | 1.045 | 0.217 | 0.072 | 0.016 | 0.023 | 0.023 | 0.005 | 0.067 |
| 2001 | 0.000 | 0.000 | 0.002 | 0.102 | 0.161 | 0.427 | 0.039 | 0.096 | 0.296 | 0.839 | 0.507 | 0.074 | 0.024 | 0.004 | 0.003 | 0.000 | 0.022 |
| 2002 | 0.000 | 0.000 | 0.062 | 0.198 | 0.643 | 0.256 | 0.326 | 0.030 | 0.094 | 0.265 | 0.663 | 0.339 | 0.053 | 0.012 | 0.007 | 0.000 | 0.010 |
| 2003 | 0.000 | 0.003 | 0.005 | 0.075 | 0.326 | 0.750 | 0.182 | 0.173 | 0.024 | 0.077 | 0.222 | 0.578 | 0.223 | 0.044 | 0.009 | 0.002 | 0.003 |
| 2004 | 0.000 | 0.002 | 0.043 | 0.030 | 0.143 | 0.878 | 1.198 | 0.196 | 0.163 | 0.033 | 0.067 | 0.185 | 0.412 | 0.210 | 0.026 | 0.005 | 0.006 |

Table 3.5.4.2 Norwegian Spring Spawning Herring Landings in numbers ('000) by length group and quarters in the Norwegian Sea 2004.

| Length (cm) | $\begin{gathered} \text { Quarter } \\ 1 \\ \hline \end{gathered}$ | $\begin{gathered} \text { Quarter } \\ 2 \\ \hline \end{gathered}$ | $\begin{gathered} \text { Quarter } \\ 3 \\ \hline \end{gathered}$ | $\begin{gathered} \text { Quarter } \\ 4 \\ \hline \end{gathered}$ | All year |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 13 |  |  |  | 59 | 59 |
| 14 |  |  |  |  |  |
| 15 |  |  |  | 59 | 59 |
| 16 |  |  |  | 59 | 59 |
| 17 |  |  |  | 59 | 59 |
| 18 | 349 | 1 |  | 118 | 468 |
| 19 |  |  | 36 | 176 | 212 |
| 20 |  |  | 71 | 176 | 247 |
| 21 |  |  | 36 | 59 | 95 |
| 22 | 217 |  | 71 | 411 | 699 |
| 23 | 1037 | 1257 | 1177 | 705 | 4175 |
| 24 | 369 | 838 | 0 | 1116 | 2323 |
| 25 | 574 | 843 | 944 | 2350 | 4711 |
| 26 | 703 | 3360 | 4171 | 5170 | 13403 |
| 27 | 2108 | 10505 | 3207 | 6992 | 22812 |
| 28 | 6090 | 55895 | 7538 | 6405 | 75928 |
| 29 | 16713 | 126468 | 14828 | 6698 | 164707 |
| 30 | 40044 | 129206 | 41312 | 14924 | 225486 |
| 31 | 97087 | 54552 | 60581 | 32844 | 245064 |
| 32 | 91173 | 21166 | 57897 | 66803 | 237039 |
| 33 | 80183 | 18351 | 50843 | 73912 | 223289 |
| 34 | 88380 | 15496 | 55894 | 73735 | 233505 |
| 35 | 130428 | 5773 | 83817 | 110869 | 330887 |
| 36 | 83426 | 3047 | 68533 | 93714 | 248720 |
| 37 | 19342 | 1029 | 25114 | 37427 | 82912 |
| 38 | 4342 | 77 | 5202 | 6110 | 15731 |
| 39 | 932 | 8 | 1158 | 1058 | 3157 |
| 40 | 245 | 2 |  | 59 | 306 |
| 41 |  |  |  | 59 | 59 |
| TOTAL numbers | 663743 | 447874 | 482428 | 542126 | 2136171 |
| Official Catch (t) | 210817 | 104857 | 167405 | 187507 | 670586 |

## Table 3.5.5.1 Weight at age in the stock for Norwegian spring spawning herring, gram

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 1 | 8 | 47 | 100 | 204 | 230 | 255 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1951 | 1 | 8 | 47 | 100 | 204 | 230 | 255 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1952 | 1 | 8 | 47 | 100 | 204 | 230 | 255 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1953 | 1 | 8 | 47 | 100 | 204 | 230 | 255 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1954 | 1 | 8 | 47 | 100 | 204 | 230 | 255 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1955 | 1 | 8 | 47 | 100 | 195 | 213 | 260 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1956 | 1 | 8 | 47 | 100 | 205 | 230 | 249 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1957 | 1 | 8 | 47 | 100 | 136 | 228 | 255 | 262 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1958 | 1 | 8 | 47 | 100 | 204 | 242 | 292 | 295 | 293 | 305 | 315 | 330 | 340 | 345 | 352 | 360 | 365 |
| 1959 | 1 | 8 | 47 | 100 | 204 | 252 | 260 | 290 | 300 | 305 | 315 | 325 | 330 | 340 | 345 | 355 | 360 |
| 1960 | 1 | 8 | 47 | 100 | 204 | 270 | 291 | 293 | 321 | 318 | 320 | 344 | 349 | 370 | 379 | 375 | 380 |
| 1961 | 1 | 8 | 47 | 100 | 232 | 250 | 292 | 302 | 304 | 323 | 322 | 321 | 344 | 357 | 363 | 365 | 370 |
| 1962 | 1 | 8 | 47 | 100 | 219 | 291 | 300 | 316 | 324 | 326 | 335 | 338 | 334 | 347 | 354 | 358 | 358 |
| 1963 | 1 | 8 | 47 | 100 | 185 | 253 | 294 | 312 | 329 | 327 | 334 | 341 | 349 | 341 | 358 | 375 | 375 |
| 1964 | 1 | 8 | 47 | 100 | 194 | 213 | 264 | 317 | 363 | 353 | 349 | 354 | 357 | 359 | 365 | 402 | 402 |
| 1965 | 1 | 8 | 47 | 100 | 186 | 199 | 236 | 260 | 363 | 350 | 370 | 360 | 378 | 387 | 390 | 394 | 394 |
| 1966 | 1 | 8 | 47 | 100 | 185 | 219 | 222 | 249 | 306 | 354 | 377 | 391 | 379 | 378 | 361 | 383 | 383 |
| 1967 | 1 | 8 | 47 | 100 | 180 | 228 | 269 | 270 | 294 | 324 | 420 | 430 | 366 | 368 | 433 | 414 | 414 |
| 1968 | 1 | 8 | 47 | 100 | 115 | 206 | 266 | 275 | 274 | 285 | 350 | 325 | 363 | 408 | 388 | 378 | 378 |
| 1969 | 1 | 8 | 47 | 100 | 115 | 145 | 270 | 300 | 306 | 308 | 318 | 340 | 368 | 360 | 393 | 397 | 397 |
| 1970 | 1 | 8 | 47 | 100 | 209 | 272 | 230 | 295 | 317 | 323 | 325 | 329 | 380 | 370 | 380 | 391 | 391 |
| 1971 | 1 | 15 | 80 | 100 | 190 | 225 | 250 | 275 | 290 | 310 | 325 | 335 | 345 | 355 | 365 | 390 | 390 |
| 1972 | 1 | 10 | 70 | 150 | 150 | 140 | 210 | 240 | 270 | 300 | 325 | 335 | 345 | 355 | 365 | 390 | 390 |
| 1973 | 1 | 10 | 85 | 170 | 259 | 342 | 384 | 409 | 404 | 461 | 520 | 534 | 500 | 500 | 500 | 500 | 500 |
| 1974 | 1 | 10 | 85 | 170 | 259 | 342 | 384 | 409 | 444 | 461 | 520 | 543 | 482 | 482 | 482 | 482 | 482 |
| 1975 | 1 | 10 | 85 | 181 | 259 | 342 | 384 | 409 | 444 | 461 | 520 | 543 | 482 | 482 | 482 | 482 | 482 |
| 1976 | 1 | 10 | 85 | 181 | 259 | 342 | 384 | 409 | 444 | 461 | 520 | 543 | 482 | 482 | 482 | 482 | 482 |
| 1977 | 1 | 10 | 85 | 181 | 259 | 343 | 384 | 409 | 444 | 461 | 520 | 543 | 482 | 482 | 482 | 482 | 482 |
| 1978 | 1 | 10 | 85 | 180 | 294 | 326 | 371 | 409 | 461 | 476 | 520 | 543 | 500 | 500 | 500 | 500 | 500 |
| 1979 | 1 | 10 | 85 | 178 | 232 | 359 | 385 | 420 | 444 | 505 | 520 | 551 | 500 | 500 | 500 | 500 | 500 |
| 1980 | 1 | 10 | 85 | 175 | 283 | 347 | 402 | 421 | 465 | 465 | 520 | 534 | 500 | 500 | 500 | 500 | 500 |
| 1981 | 1 | 10 | 85 | 170 | 224 | 336 | 378 | 387 | 408 | 397 | 520 | 543 | 512 | 512 | 512 | 512 | 512 |
| 1982 | 1 | 10 | 85 | 170 | 204 | 303 | 355 | 383 | 395 | 413 | 453 | 468 | 506 | 506 | 506 | 506 | 506 |
| 1983 | 1 | 10 | 85 | 155 | 249 | 304 | 368 | 404 | 424 | 437 | 436 | 493 | 495 | 495 | 495 | 495 | 495 |
| 1984 | 1 | 10 | 85 | 140 | 204 | 295 | 338 | 376 | 395 | 407 | 413 | 422 | 437 | 437 | 437 | 437 | 437 |
| 1985 | 1 | 10 | 85 | 148 | 234 | 265 | 312 | 346 | 370 | 395 | 397 | 428 | 428 | 428 | 428 | 428 | 428 |
| 1986 | 1 | 10 | 85 | 54 | 206 | 265 | 289 | 339 | 368 | 391 | 382 | 388 | 395 | 395 | 395 | 395 | 395 |
| 1987 | 1 | 10 | 55 | 90 | 143 | 241 | 279 | 299 | 316 | 342 | 343 | 362 | 376 | 376 | 376 | 376 | 376 |

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| 1988 | 1 | 15 | 50 | 98 | 135 | 197 | 277 | 315 | 339 | 343 | 359 | 365 | 376 | 376 | 376 | 376 | 376 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1989 | 1 | 15 | 100 | 154 | 175 | 209 | 252 | 305 | 367 | 377 | 359 | 395 | 396 | 396 | 396 | 396 | 396 |
| 1990 | 1 | 8 | 48 | 219 | 198 | 258 | 288 | 309 | 428 | 370 | 403 | 387 | 440 | 440 | 440 | 440 | 440 |
| 1991 | 1 | 11 | 37 | 147 | 210 | 244 | 300 | 324 | 336 | 343 | 382 | 366 | 425 | 425 | 425 | 425 | 425 |
| 1992 | 1 | 7 | 30 | 128 | 224 | 296 | 327 | 355 | 345 | 367 | 341 | 361 | 430 | 470 | 470 | 470 | 450 |
| 1993 | 1 | 8 | 25 | 81 | 201 | 265 | 323 | 354 | 358 | 381 | 369 | 396 | 393 | 374 | 403 | 400 | 400 |
| 1994 | 1 | 10 | 25 | 75 | 151 | 254 | 318 | 371 | 347 | 412 | 382 | 407 | 410 | 410 | 410 | 410 | 410 |
| 1995 | 1 | 18 | 25 | 66 | 138 | 230 | 296 | 346 | 388 | 363 | 409 | 414 | 422 | 410 | 410 | 405 | 447 |
| 1996 | 1 | 18 | 25 | 76 | 118 | 188 | 261 | 316 | 346 | 374 | 390 | 390 | 384 | 398 | 398 | 398 | 398 |
| 1997 | 1 | 18 | 25 | 96 | 118 | 174 | 229 | 286 | 323 | 370 | 378 | 386 | 360 | 393 | 391 | 391 | 391 |
| 1998 | 1 | 18 | 25 | 74 | 147 | 174 | 217 | 242 | 278 | 304 | 310 | 359 | 340 | 344 | 385 | 363 | 375 |
| 1999 | 1 | 18 | 25 | 102 | 150 | 223 | 240 | 264 | 283 | 315 | 345 | 386 | 386 | 386 | 382 | 382 | 407 |
| 2000 | 1 | 18 | 25 | 102 | 150 | 223 | 240 | 264 | 283 | 315 | 345 | 386 | 386 | 386 | 382 | 382 | 407 |
| 2001 | 1 | 18 | 25 | 75 | 178 | 238 | 247 | 296 | 307 | 314 | 328 | 351 | 376 | 406 | 414 | 425 | 425 |
| 2002 | 1 | 10 | 23 | 57 | 177 | 241 | 275 | 302 | 311 | 314 | 328 | 341 | 372 | 405 | 415 | 467 | 409 |
| 2003 | 1 | 10 | 55 | 98 | 159 | 211 | 272 | 305 | 292 | 331 | 337 | 347 | 356 | 381 | 414 | 425 | 441 |
| 2004 | 1 | 10 | 55 | 106 | 149 | 212 | 241 | 279 | 302 | 337 | 354 | 355 | 360 | 371 | 400 | 412 | 445 |
| 2005 | 1 | 10 | 46 | 112 | 156 | 234 | 267 | 295 | 330 | 363 | 377 | 414 | 406 | 308 | 420 | 452 | 452 |

## Table 3.5.5.2 Norwegian spring spawning herring. Catch weight at age (in kg).

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 1 | 12 | 13 | 14 | 15 | + |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0.007 | 0.025 | 0.058 | 0.11 | 0.188 | 0.211 | 0.234 | 0.253 | 0.266 | 0.28 | 0.294 | 0.303 | 0.312 | 0.32 | 0.323 | 0.331 | 0.335 |
| 1951 | 0.009 | 0.029 | 0.068 | 0.13 | 0.222 | 0.249 | 0.276 | 0.298 | 0.314 | 0.33 | 0.346 | 0.357 | 0.368 | 0.377 | 0.381 | 0.39 | 0.395 |
| 1952 | 0.008 | 0.026 | 0.061 | 0.115 | 0.197 | 0.221 | 0.245 | 0.265 | 0.279 | 0.293 | 0.308 | 0.317 | 0.327 | 0.335 | 0.339 | 0.346 | 0.351 |
| 1953 | 0.008 | 0.027 | 0.063 | 0.12 | 0.205 | 0.2 | 0.255 | 0.275 | 0. | 0.305 | 0.32 | 0.33 | 0. | 0.347 | 0.351 | 0.359 | 0.364 |
| 1954 | 0.008 | 0.026 | 0.062 | 0.117 | 0.201 | 0.225 | 0.25 | 0.269 | 0.284 | 0.299 | 0.313 | 0.323 | 0.333 | 0.341 | 0.345 | 0.352 | 0.357 |
| 1955 | 0.008 | 0.027 | 0.063 | 0.119 | 0. | 0. | 0.254 | 0.274 | 0.289 | 0.304 | 0.318 | 0.328 | 0.338 | 0.346 | 0.35 | 0.358 | 0.363 |
| 1956 | 0.008 | 0.028 | 0.066 | 0.126 | 0. | 0. | 0.268 | 0.289 | 0.304 | 0.32 | 0.336 | 0.346 | 0.357 | 0.365 | 0.369 | 0.378 | 0.383 |
| 1957 | 0.008 | 0.028 | 0.066 | 0 . | 0. | 0. | 0. | 0. | 0.306 | 0.322 | 0.338 | 0.348 | 0.359 | 0.367 | 0.371 | 0.38 | 0.385 |
| 1958 | 0.009 | 0.0 | 0.0 | 0.133 | 0. | 0. | 0. | 0.305 | 0 | 0.338 | 0.355 | 0.366 | 0.377 | 0.386 | 0.39 | 0.399 | 0.404 |
| 1959 | 0.009 | 0.03 | 0.0 | 0.135 | 0. | 0. | 0. | 0. | 0. | 0.344 | 0.36 | 0.372 | 0.383 | 0.392 | 0.397 | 0.406 | 0.411 |
| 1960 | 0.0 | 0.011 | 0. | 0.119 | 0. |  | 0 | 0.318 | 0.363 | 0.379 | 0. | 0 | 0.411 | 0.439 | 0.45 | 0.444 | 0.448 |
| 19 | 0.0 | 0. | 0. | 0.087 | 0. | 0. | 0 | 0 | 0 | 0.393 | 0. | 0. | 0.422 | 0.447 | 0.465 | 0.452 | 0.452 |
| 19 | 0.0 | 0.0 | 0. | 0. | 0. | 0. | 0. | 0 | 0 | 0. | 0. | 0. | 0. | 0.394 | 0.399 | 0.411 | 0.416 |
| 1963 | 0.008 | 0.0 | 0. | 0.098 | 0. | 0. | 0. | 0. | 0 | 0.33 | 0.341 | 0.358 | 0.385 | 0.353 | 0.381 | 0.386 | 0.386 |
| 1964 | 0.0 | 0.0 | 0. | 0 | 0. | 0. | 0. | 0. | 0. | 0 | 0. | 0.351 | 0.367 | 0.375 | 0.372 | 0.427 | 0.434 |
| 1965 | 0.0 | 0.0 | 0.0 | 0.089 | 0. | . | 0 | 0. | 0 | 0. | 0. | 0.395 | 0.393 | 0.404 | 0.401 | 0.429 | 0.437 |
| 19 | 0.0 | 0.0 | 0. | 0.063 | 0.2 | 0. | 0 | 0 | 0 | 0.425 | 0. | 0 | 0.467 | 0.446 | 0.459 | 0.465 | 0.474 |
| 1967 | 0.009 | 0.015 | 0. | 0. | 0. | 0. | 0. | 0.31 | 0.333 | 0.359 | 0 . | 0.446 | 0.401 | 0.408 | 0.439 | 0.427 | 0.431 |
| 1968 | 0.01 | 0.027 | 0. | 0. | 0.108 | 0 | 0. | 0.38 | 0.364 | 0.382 | 0 | 0. | 0.442 | 0.517 | 0.491 | 0.464 | 0.487 |
| 1969 | 0.009 | 0.021 | 0.0 | 0. | 0. | 0 | 0. | 0. | 0 | 0. | 0. | 0.363 | 0.385 | 0.377 | 0 | 0.423 | 0.429 |
| 1970 | 0.008 | 0.058 | 0.085 | 0. | 0. | 0.25 | 0 | 0.2 | 0.298 | 0. | 0. | 0.309 | 0.357 | 0.348 | 0.357 | 0.367 | 0.376 |
| 1971 | 0.011 | 0.053 | 0.121 | 0.177 | 0.2 | 0.25 | 0. | 0.305 | 0.333 | 0.353 | 0.3 | 0.377 | 0.388 | 0.399 | 0.419 | 0.444 | 0.444 |
| 1972 | 0.011 | 0.029 | 0.062 | 0.103 | 0. | 0. | 0.258 | 0.295 | 0.322 | 0.341 | 0.354 | 0.365 | 0.376 | 0.387 | 0.406 | 0.43 | 0.43 |
| 1973 | 0.006 | 0.053 | 0.106 | 0.161 | 0.213 | 0. | 0.255 | 0. | 0.287 | 0.324 | 0.338 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 | 0.257 |
| 1974 | 0.006 | 0.055 | 0. | 0.168 | 0.222 | 0.2 | 0.265 | 0.2 | 0.299 | 0.337 | 0.352 | 0.267 | 0.324 | 0.324 | 0.324 | 0.324 | 0.324 |
| 1975 | 0.009 | 0.079 | 0.169 | 0.241 | 0.318 | 0.358 | 0. | 0.413 | 0.429 | 0.484 | 0.506 | 0.384 | 0.466 | 0.466 | 0.466 | 0.466 | 0.466 |
| 1976 | 0.007 | 0.062 | 0.132 | 0.189 | 0.25 | 0.28 | 0.298 | 0.323 | 0.336 | 0.379 | 0.396 | 0.3 | 0.364 | 0.364 | 0.364 | 0.364 | 0.364 |
| 1977 | 0.011 | 0.091 | 0.193 | 0.316 | 0.35 | 0.398 | 0.439 | 0.495 | 0.511 | 0.558 | 0.583 | 0.537 | 0.537 | 0.537 | 0.537 | 0.537 | 0.537 |
| 1978 | 0.012 | 0.1 | 0.21 | 0.274 | 0.424 | 0.454 | 0.495 | 0.524 | 0.596 | 0.613 | 0.65 | 0.59 | 0.59 | 0.59 | 0.59 | 0.59 | 0.59 |
| 1979 | 0.01 | 0.088 | 0.181 | 0.293 | 0.359 | 0.416 | 0.436 | 0.482 | 0.482 | 0.539 | 0.553 | 0.518 | 0.518 | 0.518 | 0.518 | 0.518 | 0.518 |
| 1980 | 0.012 | 0.101 | 0.202 | 0.266 | 0.399 | 0.449 | 0.46 | 0.485 | 0.472 | 0.618 | 0.645 | 0.608 | 0.594 | 0.594 | 0.594 | 0.594 | 0.594 |

## Table 3.5.5.2 Norwegian spring spawning herring. Catch weight at age (in kg) (con't).

| 1981 | 0.01 | 0.082 | 0.163 | 0.196 | 0.291 | 0.341 | 0.368 | 0.38 | 0.397 | 0.436 | 0.45 | 0.492 | 0.481 | 0.481 | 0.481 | 0.481 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 982 | 0.01 | 0.087 | 0.159 | 0.256 | 0.312 | 0.378 | 0.415 | 0.435 | 0.449 | 0.448 | 0.506 | 0.493 | 0.499 | 0.499 | 0.499 | 0.499 | 0.499 |
| 983 | 0.011 | 0. | 0 | 0 | 0.265 | 0 | 0 | 0.41 | 0.426 | 0.435 |  | 0.468 |  |  |  | 0.461 | 0.461 |
| 1984 | 0.009 | 0 | 0 | 0 | 0.262 | 0.325 |  | 0.381 | 0.4 | 0.413 | 0.405 |  | 0.415 | 0.415 | 0.415 | 0.415 | 15 |
| 1 | 0.009 | 0.022 | 0 | 0 | 0.277 | 0 | 0 | 0.36 | 0.381 | 0.397 | 9 | 0 | 0 | 0.435 | 0.435 | 5 | 35 |
| 1 | 0 | 0.077 | 0 | 0 | 0 | 0.294 | 0.312 | 0.352 | 0.374 | 0.398 |  | 0.401 | 0 | 0.41 | 0.41 | 0.41 | 0.41 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.312 | 0 | - | 0.353 | 0.37 | 0 | 0.385 | 5 | 0.385 | 85 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0.241 | 0.265 | O | 0 | 0.317 | 0.308 | 0.334 | 4 | 0.334 | 4 | 34 |
| 1 | 0. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | - | 0.422 | 0 | 0.429 | 9 | 0.429 | 0.429 | 9 |
| 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.308 | 0 | O | 2 | O | 0.428 | 8 | 8 | 8 | 28 |
| 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.316 | 0 | - | 2 | 0.354 | 0.398 | 8 | 8 | 0.398 | 8 |
| 19 | 0. | 0 | 0 |  | 0 | 0 | 0 | 0. | 0 |  | 0.403 | 0.365 | 0. | 4 | 6 | 0 | 0.41 |
| 19 | 0. | 0 | 0 |  | 0 | 0 | 0 | 0 | O | 0.373 | 0.379 | 0 | 0 | 0.39 | 0.395 | 0 | 05 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.372 | 0 | 0 | 0 | 03 | 405 | 407 | 0.405 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.356 | 0.374 | 0.366 | 0 | 0 | 0 | 0 | 0.4 |
| 1 | 0 | 0. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.366 | 0.36 | 0.361 | 0 | 0.379 | 0.379 | 0.379 | 379 |
| 1 | 0 | 0. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.328 | 0.344 | 0.343 | 0.397 | 0.357 | 0.51 | 0.51 |
| 1 | 0. | 0. | 0 | 0 | 0 | 0 | 0 | 0 | 0.316 | 0.338 | 0.374 | 0.372 | 0.366 | 0.396 | 0.377 | 0.406 | 0.406 |
| 1999 | 0.0 | 0. | 0. | 0 | 0.228 | 0. | 0. | 0.292 | 0.307 | 0.335 | 0.362 | 0.371 | 0.399 | 0.396 | 0 | 0. | 0.404 |
| 2000 | 0.0 | 0. | 0. | 0.1 | 0.222 | 0.242 | 0.2 | 0.303 | 0.31 | 0.328 | 0.349 | 0.383 | 0.411 | 0.41 | 0.419 | 0.409 | 0.409 |
| 20 | 0.007 | 0.0 | 0.1 | 0. | 0.214 | 0.252 | 0.268 | 0.305 | 0.308 | 0.322 | 0.337 | 0.363 | 0.353 | 0.378 | 0. | 0.427 | 0.427 |
| 2002 | 0.007 | 0. | 0.0 | 0.128 | 0.198 | 0.255 | 0.28 | 0.303 | 0.322 | 0.323 | 0.334 | 0.345 | 0.369 | 0.407 | 0.41 | 0.435 | 0.435 |
| 2003 | 0.007 | 0.0 | 0.068 | 0.17 | 0.216 | 0.255 | 0.286 | 0.317 | 0.325 | 0.351 | 0.36 | 0.358 | 0.378 | 0.394 | 0.399 | 0.447 | 0.447 |
| 2004 | 0.007 | 0.065 | 0.143 | 0.183 | 0.211 | 0.236 | 0.263 | 0.297 | 0.342 | 0.358 | 0.361 | 0.382 | 0.389 | 0.399 | 0.42 | 0.439 | 0.439 |

## Table 3.5.7.1 Proportion mature at age - Norwegian spring spawning herring.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 4 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0.00 | 0.00 | 0.00 | 0.00 | 0.10 | 0.30 | 0.60 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1951 | 0.00 | 0.00 | 0.00 | 0.00 | 0.10 | 0.30 | 0.60 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1952 | 0.00 | 0.00 | 0.00 | 0.00 | 0.10 | 0.30 | 0.60 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1953 | 0.00 | 0.00 | 0.00 | 0.00 | 0.10 | 0.30 | 0.60 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1954 | 0.00 | 0.00 | 0.00 | 0.00 | 0.10 | 0.30 | 0.60 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1955 | 0.00 | 0.00 | 0.00 | 0.08 | 0.22 | 0.37 | 0.85 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1956 | 0.00 | 0.00 | 0.00 | 0.08 | 0.22 | 0.37 | 0.85 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1957 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.50 | 0.60 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1958 | 0.00 | 0.00 | 0.00 | 0.08 | 0.22 | 0.37 | 0.85 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1959 | 0.00 | 0.00 | 0.00 | 0.08 | 0.22 | 0.37 | 0.85 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1960 | 0.00 | 0.00 | 0.00 | 0.08 | 0.22 | 0.37 | 0.85 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1961 | 0.00 | 0.00 | 0.00 | 0.04 | 0.35 | 0.68 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1962 | 0.00 | 0.00 | 0.00 | 0.00 | 0.11 | 0.67 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1963 | 0.00 | 0.00 | 0.00 | 0.04 | 0.03 | 0.32 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1964 | 0.00 | 0.00 | 0.00 | 0.02 | 0.06 | 0.28 | 0.32 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1965 | 0.00 | 0.00 | 0.00 | 0.00 | 0.34 | 0.35 | 0.76 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1966 | 0.00 | 0.00 | 0.00 | 0.01 | 0.15 | 1.00 | 0.96 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1967 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.23 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1968 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.76 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1969 | 0.00 | 0.00 | 0.00 | 0.62 | 0.89 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1970 | 0.00 | 0.00 | 0.00 | 0.06 | 0.13 | 0.31 | 0.17 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1971 | 0.00 | 0.00 | 0.00 | 0.10 | 0.25 | 0.60 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1972 | 0.00 | 0.00 | 0.00 | 0.00 | 0.10 | 0.25 | 0.60 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1973 | 0.00 | 0.00 | 0.00 | 0.50 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1974 | 0.00 | 0.00 | 0.00 | 0.50 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1975 | 0.00 | 0.00 | 0.00 | 0.50 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1976 | 0.00 | 0.00 | 0.00 | 0.50 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1977 | 0.00 | 0.00 | 0.00 | 0.73 | 0.89 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1978 | 0.00 | 0.00 | 0.00 | 0.13 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1979 | 0.00 | 0.00 | 0.00 | 0.10 | 0.62 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1980 | 0.00 | 0.00 | 0.00 | 0.25 | 0.50 | 0.97 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1981 | 0.00 | 0.00 | 0.00 | 0.30 | 0.50 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1982 | 0.00 | 0.00 | 0.00 | 0.10 | 0.48 | 0.70 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1983 | 0.00 | 0.00 | 0.00 | 0.10 | 0.50 | 0.69 | 0.71 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1984 | 0.00 | 0.00 | 0.00 | 0.10 | 0.50 | 0.90 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1985 | 0.00 | 0.00 | 0.00 | 0.10 | 0.50 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1986 | 0.00 | 0.00 | 0.00 | 0.10 | 0.20 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1987 | 0.00 | 0.00 | 0.00 | 0.10 | 0.30 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |

Table 3.5.7.1 Proportion mature at age - Norwegian spring spawning herring (con't).

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1988 | 0.00 | 0.00 | 0.00 | 0.10 | 0.30 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1989 | 0.00 | 0.00 | 0.00 | 0.10 | 0.30 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1990 | 0.00 | 0.00 | 0.00 | 0.40 | 0.80 | 0.90 | 0.90 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1991 | 0.00 | 0.00 | 0.00 | 0.10 | 0.70 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1992 | 0.00 | 0.00 | 0.00 | 0.10 | 0.20 | 0.80 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1993 | 0.00 | 0.00 | 0.00 | 0.01 | 0.30 | 0.80 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1994 | 0.00 | 0.00 | 0.00 | 0.01 | 0.30 | 0.80 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1995 | 0.00 | 0.00 | 0.00 | 0.00 | 0.30 | 0.80 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1996 | 0.00 | 0.00 | 0.00 | 0.00 | 0.30 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1997 | 0.00 | 0.00 | 0.00 | 0.00 | 0.30 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1998 | 0.00 | 0.00 | 0.00 | 0.00 | 0.30 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1999 | 0.00 | 0.00 | 0.00 | 0.00 | 0.30 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2000 | 0.00 | 0.00 | 0.00 | 0.00 | 0.30 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2001 | 0.00 | 0.00 | 0.00 | 0.00 | 0.30 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2002 | 0.00 | 0.00 | 0.00 | 0.00 | 0.30 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2003 | 0.00 | 0.00 | 0.00 | 0.00 | 0.30 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2004 | 0.00 | 0.00 | 0.00 | 0.00 | 0.30 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2005 | 0.00 | 0.00 | 0.00 | 0.10 | 0.30 | 0.90 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 200 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 3.6.1.1.1 Norwegian Spring-spawning herring. Estimates from the acoustic surveys on the spawning stock in February-March. Numbers in millions.

| Year <br> Age | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2005* |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 |  | 101 | 183 | 44 |  |  | 16 |  | 407 |  |  | 106 | 1516 | 103 |
| 3 | 255 | 5 | 187 | 59 |  |  | 128 | 1792 | 231 |  |  | 1366 | 690 | 281 |
| 4 | 146 | 373 | 0 | 54 |  |  | 676 | 7621 | 7638 |  | 381 | 337 | 1996 | 811 |
| 5 | 6805 | 103 | 345 | 12 |  |  | 1375 | 3807 | 11243 |  | 1905 | 1286 | 164 | 3310 |
| 6 | 202 | 5402 | 112 | 354 |  |  | 476 | 2151 | 2586 |  | 10640 | 2979 | 592 | 7545 |
| 7 |  | 182 | 4489 | 122 |  |  | 63 | 322 | 957 |  | 6708 | 11791 | 1997 | 10453 |
| 8 |  |  | 146 | 4148 |  |  | 13 | 20 | 471 |  | 1280 | 7534 | 7714 | 887 |
| 9 |  |  |  | 102 |  |  | 140 | 1 | 0 |  | 434 | 1912 | 4240 | 563 |
| 10 |  |  |  |  |  |  | 35 | 124 | 0 |  | 130 | 568 | 553 | 159 |
| 11 |  |  |  |  |  |  | 1820 | 63 | 165 |  | 39 | 132 | 71 | 122 |
| 12 |  |  |  |  |  |  |  | 2573 | 0 |  | 0 | 0 | 3 | 610 |
| 13 |  |  |  |  |  |  |  |  | 2024 |  | 175 | 0 | 0 | 1100 |
| 14 |  |  |  |  |  |  |  |  |  |  | 0 | 392 | 6 | 686 |
| 15+ |  |  |  |  |  |  |  |  |  |  | 804 | 437 | 361 | 17 |
| Total | 7408 | 6166 | 5462 | 4895 | - | - | 4742 | 18474 | 25756 | - | 22496 | 28840 | 19903 | 26649 |

In 1992, 1993 and 1997 there was no estimate due to poor weather conditions.

* No surveys in 2001-2004.

Table 3.6.1.2.1 Norwegian Spring-spawning herring. Estimates obtained on the acoustic surveys in the wintering areas in November-December. Numbers in millions.

| Year <br> Age | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2 *}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 |  | 72 |  | 380 |  | 9 | 65 | 74 | 56 | 362 | 7 |
| 2 | 36 | 1518 | 16 | 183 | 1465 | 73 | 1207 | 159 | 322 | 522 | 50 |
| 3 | 1247 | 2389 | 3708 | 5133 | 3008 | 661 | 441 | 2425 | 1522 | 3916 | 276 |
| 4 | 1317 | 3287 | 4124 | 5274 | 13180 | 1480 | 1833 | 296 | 5260 | 1528 | 1659 |
| 5 | 173 | 1267 | 2593 | 1839 | 5637 | 6110 | 3869 | 837 | 165 | 2615 | 624 |
| 6 | 16 | 13 | 1096 | 1040 | 994 | 4458 | 12052 | 2066 | 497 | 82 | 1029 |
| 7 | 208 | 13 | 34 | 308 | 552 | 1843 | 8242 | 6601 | 1869 | 338 | 32 |
| 8 | 139 | 158 | 25 | 19 | 92 | 743 | 2068 | 4168 | 4785 | 864 | 188 |
| 9 | 3742 | 26 | 196 | 13 | 0 | 66 | 629 | 755 | 3635 | 3160 | 516 |
| 10 | 69 | 4435 | 29 | 111 | 7 | 0 | 111 | 212 | 668 | 2216 | 1831 |
| 11 |  |  | 3239 | 39 | 41 | 0 | 14 | 0 | 205 | 384 | 911 |
| 12 |  |  |  | 907 | 15 | 126 | 0 | 15 | 0 | 127 | 184 |
| 13 |  |  |  |  | 393 | 0 | 392 | 0 | 0 | 0 | 0 |
| $14+$ |  |  |  |  |  | 842 | 221 | 146 | 168 | 18 | 0 |
| Total | 6947 | 13178 | 15209 | 15246 | 25384 | 16411 | 31144 | 17754 | 19152 | 16132 | 7345 |


| Year <br> Age | 2003** | 2004** |
| :--- | ---: | ---: |
| 1 | 586 | 257 |
| 2 | 406 | 6814 |
| 3 | 2167 | 1123 |
| 4 | 10670 | 1596 |
| 5 | 13237 | 5334 |
| 6 | 1047 | 6731 |
| 7 | 678 | 363 |
| 8 | 41 | 280 |
| 9 | 134 | 37 |
| 10 | 301 | 42 |
| 11 | 1214 | 187 |
| 12 | 502 | 761 |
| 13 | 10 | 392 |
| $14+$ | 37 | 83 |
| Total | 31030 | 24000 |

* Much of the youngest yearclasses ( $-98,-99$ ) wintered outside the fjords this winter and are not included in the estimate
** In 2003-2004 a combined estimate from the Tysfjord, Ofotfjord and oceanic areas off Vesterålen/Troms.

Table 3.6.1.2.2 Norwegian Spring Spawning herring. Estimates obtained on the acoustic surveys in the wintering areas in January. Numbers in millions. No surveys carried out since 1999.

| Year <br> Age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 90 |  |  |  | 73 |  |  |  | 214 |
| 3 | 220 | 410 | 61 | 642 | 47 | 315 |  | 267 | 1358 |
| 4 | 70 | 820 | 1905 | 3431 | 3781 | 10442 |  | 1938 | 199 |
| 5 | 20 | 260 | 2048 | 4847 | 4013 | 13557 |  | 4162 | 1455 |
| 6 | 180 | 60 | 256 | 1503 | 2445 | 4312 |  | 9647 | 4452 |
| 7 | 150 | 510 | 27 | 102 | 1215 | 1271 |  | 6974 | 12971 |
| 8 | 5500 | 120 | 269 | 29 | 42 | 290 |  | 1518 | 7226 |
| 9 | 440 | 4690 | 182 | 161 | 24 | 22 |  | 743 | 1876 |
| 10 |  | 30 | 5691 | 131 | 267 | 25 |  | 16 | 499 |
| 11 |  |  | 128 | 3679 | 29 | 200 |  | 4 | 16 |
| 12 |  |  |  |  | 4326 | 58 |  | 0 | 16 |
| 13 |  |  |  |  |  | 1146 |  | 181 | 0 |
| 14 |  |  |  |  |  |  |  | 7 | 156 |
| $15+$ |  |  |  |  |  |  |  | 314 | 220 |
| Total | 6670 | 6900 | 10567 | 14598 | 16189 | 31638 | - | 25985 | 30444 |

In 1997 there was no estimate due to poor weather conditions.
Table 3.6.1.3.1 Norwegian spring-spawning herring. Estimates from the international acoustic surveys on the feeding areas in the Norwegian Sea in May. Numbers in millions.

| $\begin{aligned} & \text { Year } \\ & \text { Age } \\ & \hline \end{aligned}$ | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0 | 0 | 24 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 1404 | 215 | 157 | 1540 | 677 |
| 3 | 4114 | 1169 | 367 | 2191 | 1353 | 8312 | 6343 |
| 4 | 22461 | 3599 | 1099 | 322 | 2783 | 1430 | 9619 |
| 5 | 13244 | 18867 | 4410 | 965 | 92 | 1463 | 1418 |
| 6 | 4916 | 13546 | 16378 | 3067 | 384 | 179 | 779 |
| 7 | 2045 | 2473 | 10160 | 11763 | 1302 | 204 | 375 |
| 8 | 424 | 1771 | 2059 | 6077 | 7194 | 3215 | 847 |
| 9 | 14 | 178 | 804 | 853 | 5344 | 5433 | 1941 |
| 10 | 7 | 77 | 183 | 258 | 1689 | 1220 | 2500 |
| 11 | 155 | 288 | 0 | 5 | 271 | 94 | 1423 |
| 12 | 0 | 415 | 0 | 14 | 0 | 178 | 61 |
| 13 | 3134 | 60 | 112 | 0 | 114 | 0 | 78 |
| 14 |  | 2472 | 0 | 158 | 0 | 0 | 28 |
| 15+ |  |  | 415 | 128 | 1135 | 85 | 26 |
| Total | 50514 | 44915 | 37415 | 26016 | 21857 | 23353 | 26142 |
| Year | 2003 | 2004 | 2005 |  |  |  |  |
| Age |  |  |  |  |  |  |  |
| 1 | 32073 | 0 | 0 |  |  |  |  |
| 2 | 8115 | 13735 | 1293 |  |  |  |  |
| 3 | 6561 | 1543 | 19679 |  |  |  |  |
| 4 | 9985 | 5227 | 1353 |  |  |  |  |
| 5 | 9961 | 12571 | 1765 |  |  |  |  |
| 6 | 1499 | 10710 | 6205 |  |  |  |  |
| 7 | 732 | 1075 | 5371 |  |  |  |  |
| 8 | 146 | 580 | 651 |  |  |  |  |
| 9 | 228 | 76 | 388 |  |  |  |  |
| 10 | 1865 | 313 | 139 |  |  |  |  |
| 11 | 2359 | 362 | 262 |  |  |  |  |
| 12 | 1769 | 1294 | 526 |  |  |  |  |
| 13 |  | 1120 | 1003 |  |  |  |  |
| 14 | 287 | 10 | 364 |  |  |  |  |
| 15+ | 45 | 88 | 115 |  |  |  |  |
| Total | 75625 | 48704 | 39114 |  |  |  |  |

 footnotes.

| Year | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}^{1}$ | $\mathbf{1 9 9 7}^{2}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 5}^{*}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 24.3 | 32.6 | 102.7 | 6.6 | 0.5 | 0.1 | 2.6 | 9.5 | 49.5 | 105.4 | 0.3 | 0.5 | 23.3 |
| 2 | 5.2 | 14.0 | 25.8 | 59.2 | 7.7 | 0.25 | 0.04 | 4.7 | 4.9 | 27.9 | 7.6 | 3.9 | 4.5 |
| 3 |  | 5.7 | 1.5 | 18.0 | 8.0 | 1.8 | 0.4 | 0.01 | 0.00 | 0.00 | 8.8 | 0.0 | 2.5 |
| 4 |  |  |  | 1.7 | 1.1 | 0.6 | 0.35 | 0.01 | 0.00 | 0.00 | 0.00 | 0.0 | 0.4 |
| 5 |  |  |  |  |  | 0.03 | 0.05 | 0.00 | 0.00 | 0.00 | 0.00 | 0.0 | 0.3 |

* No surveys in 2003 and 2004.
 3.6.1.4.4).

| Year <br> Age | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 1 | 14.7 | 0.5 | 1,3 | 99.9 | 14.3 |
| 2 | 11.5 | 10.5 | 0.0 | 4.3 | 36.5 |
| 3 | 0.0 | 1.7 | 0.0 | 2.5 | 0.9 |

Table 3.6.1.4.3 Norwegian spring spawners. Acoustic abundance ( $\mathbf{T S}=20 \operatorname{logL}$ - 71.9) of 0-group herring in Norwegian coastal waters in 1975-2004 (numbers in millions).

| Year | Area |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | South of $62{ }^{\circ} \mathrm{N}$ | $62^{\circ} \mathrm{N}-65^{\circ} \mathrm{N}$ | $65^{\circ} \mathrm{N}-68^{\circ} \mathrm{N}$ | North of $68^{\circ} 30^{\prime} \mathrm{N}$ |  |
| 1975 |  | 164 | 346 | 28 | 538 |
| 1976 |  | 208 | 1305 | 375 | 1888 |
| 1977 |  | 35 | 153 | 19 | 207 |
| 1978 |  | 151 | 256 | 196 | 603 |
| 1979 |  | 455 | 1130 | 144 | 1729 |
| 1980 |  | 6 | 2 | 109 | 117 |
| 1981 |  | 132 | 1 | 1 | 134 |
| 1982 |  | 32 | 286 | 1151 | 1469 |
| 1983 |  | 162 | 2276 | 4432 | 6866 |
| 1984 |  | 2 | 234 | 465 | 701 |
| 1985 |  | 221 | 177 | 104 | 502 |
| 1986 |  | 5 | 72 | 127 | 204 |
| 1987 |  | 327 | 26 | 57 | 410 |
| 1988 |  | 14 | 552 | 708 | 1274 |
| 1989 |  | 575 | 263 | 2052 | 2890 |
| 1990 |  | 75 | 146 | 788 | 1009 |
| 1991 |  | 80 | 299 | 2428 | 2807 |
| 1992 | 73 | 1993 | 204 | 621 | 2891 |
| 1993 | 290 | 109 | 140 | 288 | 827 |
| 1994 | 157 | 452 | 323 | 6168 | 7101 |
| 1995 | 0 | 27 | 2 | 0 | 29 |
| 1996 | 0 | 20 | 114 | 8800 | 8934 |
| 1997 | 208 | 69 | 544 | 5244 | 6065 |
| 1998 | 424 | 273 | 442 | 11640 | 12779 |
| 1999 | 121 | 658 | 271 | 6329 | 7379 |
| 2000 | 570 | 127 | 996 | 7237 | 8930 |
| 2001 | 89 | 324 | 134 | 1421 | 1968 |
| 2002 | 67 | 1227 | 284 | 3573 | 5151 |
| * | South of $62^{\circ} \mathrm{N}$ | $62^{\circ} \mathrm{N}-64^{\circ} \mathrm{N}$ | $64^{\circ} \mathrm{N}-67^{\circ} \mathrm{N}$ | North of $67^{\circ} \mathrm{N}$ |  |
| 2003 | 9 | 44 | 6647 | 21417 | 28117 |
| 2004 | 19 | 884 | 1306 | 11950 | 14159 |

* A new survey design was introduced in 2003, which resulted in changed areas and wider and denser coverage of the fjords. Thus the estimates from 2003 and onwards are likely to be higher and not directly comparable to the former estimates.
** Data not supplied.

Table 3.6.1.4.4 Norwegian spring-spawning herring. Abundance indices for 0-group herring in the Barents Sea, 1974-2004.

| Year | Log index | Year | Log index |
| ---: | ---: | ---: | ---: |
| 1974 | 0.01 | 1989 | 0.59 |
| 1975 | 0.00 | 1990 | 0.31 |
| 1976 | 0.00 | 1991 | 1.19 |
| 1977 | 0.01 | 1992 | 1.06 |
| 1978 | 0.02 | 1993 | 0.75 |
| 1979 | 0.09 | 1994 | 0.28 |
| 1980 | - | 1995 | 0.16 |
| 1981 | 0.00 | 1996 | 0.65 |
| 1982 | 0.00 | 1997 | 0.39 |
| 1983 | 1.77 | 1998 | 0.59 |
| 1984 | 0.34 | 1999 | 0.41 |
| 1985 | 0.23 | 2000 | 0.30 |
| 1986 | 0.00 | 2001 | 0.13 |
| 1987 | 0.00 | 2002 | 0.53 |
| 1988 | 0.32 | 2003 | 0.51 |
|  |  | 2004 | 1.20 |

Table revised since 2004.
Table 3.6.1.5.1 Norwegian Spring-spawning herring. The indices for herring larvae on the Norwegian shelf for the period 1981-2003 ( $\mathrm{N} * 10^{-12}$ )

| Year | Index1 | Index 2 | Year | Index 1 | Index 2 |
| :---: | ---: | ---: | ---: | ---: | ---: |
| 1981 | 0.3 |  | 1993 | 24.7 | 78.0 |
| 1982 | 0.7 |  | 1994 | 19.5 | 48.6 |
| 1983 | 2.5 |  | 1995 | 18.2 | 36.3 |
| 1984 | 1.4 |  | 1996 | 27.7 | 81.7 |
| 1985 | 2.3 |  | 1997 | 66.6 | 147.5 |
| 1986 | 1.0 |  | 1998 | 42.4 | 138.6 |
| 1987 | 1.3 | 4.0 | 1999 | 19.9 | 73.0 |
| 1988 | 9.2 | 25.5 | 2000 | 19.8 | 127.5 |
| 1989 | 13.4 | 28.7 | 2001 | 40.7 | 131.9 |
| 1990 | 18.3 | 29.2 | 2002 | 27.1 | 113.9 |
| 1991 | 8.6 | 23.5 | 2003 | 3.7 | 18.9 |
| 1992 | 6.3 | 27.8 | 2004 | 56.4 | 175.7 |
|  |  |  | 2005 | 73.9 | $*$ |

Index 1. The total number of herring larvae found during the cruise.
Index 2. Back-calculated number of newly hatched larvae with $10 \%$ daily moratlity. The larval age is estimated from the duration of the yolksac stages and the size of the larvae.

* Not calculated in 2004.

Table 3.2.4.1 Tagging data used in the SeaStar runs. Please note that the tagging data for 2002 was considered an outlier and thus not included in the analysis - Norwegian spring spawning herring.

Tagging data for the 1983 year class

|  |  | 倣 | Recovered |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | $\begin{aligned} & \bar{\circ} \\ & \infty \\ & \stackrel{0}{0} \\ & \stackrel{0}{0} \\ & \stackrel{\tilde{y}}{0} \end{aligned}$ | $\overrightarrow{0}$ 0 0 0 0 0 0 |  |  | $\begin{aligned} & \bar{\circ} \\ & 0 \\ & 0 \\ & \stackrel{0}{0} \\ & \stackrel{\tilde{y}}{0} \end{aligned}$ |  |  |  |  |  | $\begin{aligned} & \stackrel{\rightharpoonup}{0} \\ & 0 \\ & \stackrel{\rightharpoonup}{0} \\ & \ddot{\theta} \\ & 0 \end{aligned}$ |  |  | $\begin{aligned} & \bar{\circ} \\ & 0 \\ & 0 \\ & \stackrel{\rightharpoonup}{0} \\ & \ddot{\ddot{\theta}} \\ & \end{aligned}$ | $\begin{aligned} & \bar{\circ} \\ & \infty \\ & \infty \\ & \stackrel{\rightharpoonup}{\sigma} \\ & \stackrel{\theta}{0} \\ & \stackrel{\theta}{\sigma} \end{aligned}$ |  |
| 1987 | 33067 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1988 | 38152 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1989 | 20620 | 10695 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 12 |
| 1990 | 24585 | 5489 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 10 | 4 |
| 1991 | 12558 | 5545 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 1992 | 15262 | 1737 |  |  |  |  |  |  |  |  |  |  |  |  |  | 2 |  |  | 4 |
| 1993 | 15839 | 9372 |  |  |  |  |  |  |  |  |  |  |  |  | 9 | 12 |  | 13 | 6 |
| 1994 | 5364 | 9474 |  |  |  |  |  |  |  |  |  |  | 11 |  | 4 | 8 |  | 10 | 2 |
| 1995 | 859 | 11554 |  |  |  |  |  |  |  |  |  | 7 | 9 |  | 6 | 15 |  | 10 | 6 |
| 1996 | 2879 | 4038 |  |  |  |  |  |  |  | 3 |  | 4 | 1 |  | 2 | 10 |  |  | 3 |
| 1997 | 2266 | 3867 |  |  |  |  |  |  | 0 | 0 |  | 0 | 3 |  | 2 | 3 |  |  | 0 |
| 1998 | 648 | 509 |  |  |  |  |  |  | 0 | 0 |  | 0 | 2 |  | 2 | 1 |  |  | 1 |
| 1999 |  | 379 |  |  |  | 1 | 0 |  | 0 | 0 |  | 1 | 0 |  | 0 | 1 |  |  | 0 |
| 2000 |  | 413 |  |  | 0 | 0 | 0 |  | 0 | 0 |  | 0 | 1 |  | 0 | 3 |  |  | 0 |
| 2001 |  | 35 |  |  | 0 | 0 | 0 |  | 0 | 0 |  | 0 | 0 |  | 0 | 0 |  |  | 0 |
| 2002 |  | 221 | 0 |  | 0 | 0 | 0 |  | 0 | 0 |  | 0 | 0 |  | 0 | 0 |  |  | 0 |
| 2003 |  | 0 |  |  | 0 | 0 |  |  | 0 | 0 |  | 0 | 0 |  | 0 | 0 |  |  | 0 |
| 2004 |  | 0 |  |  | 0 | 0 |  |  | 0 | 0 |  | 0 | 0 |  | 0 |  |  |  | 0 |

Tagging data for the 1984 year class


Will not be updated after 2003.

Table 3.2.4.1 continued.
Tagging data for the 1985 year class

*1985+ group

Tagging data for the 1986 year class


Will not be updated after 2003.

Table 3.2.4.1 continued
Tagging data for the 1987 year class

*1987+group

Tagging data for the 1988 year class


Table 3.2.4.1 continued
Tagging data for the 1989 year class

| $\begin{aligned} & \widehat{\sim} \\ & \underset{\sim}{\circ} \end{aligned}$ | 花 | $\stackrel{Z}{Z}$ | Recovered |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & \overrightarrow{0} \\ & \substack{0 \\ \text { ad } \\ \underset{\sim}{0} \\ \hline} \end{aligned}$ |  | $\begin{array}{\|c\|} 2000 \\ \text { release } \end{array}$ | $\begin{array}{\|c\|} 1998 \\ \text { release } \end{array}$ | $\begin{gathered} 1997 \\ \text { release } \end{gathered}$ | $\left\lvert\, \begin{gathered} 1996 \\ \text { release } \end{gathered}\right.$ | $\begin{gathered} 1995 \\ \text { release } \end{gathered}$ | $\begin{array}{\|c\|} 1994 \\ \text { release } \end{array}$ | $\left\lvert\, \begin{gathered} 1993 \\ \text { release } \end{gathered}\right.$ |
| 1993 | 7584 |  |  |  |  |  |  |  |  |
| 1994 | 11873 |  |  |  |  |  |  |  |  |
| 1995 | 2348 | 9463 |  |  |  |  |  |  |  |
| 1996 | 5170 | 4636 |  |  |  |  |  |  |  |
| 1997 | 4103 | 3346 |  |  |  |  | 0 |  | 2 |
| 1998 | 1176 | 1183 |  |  |  | 1 | 0 | 0 | 0 |
| 1999 |  | 1179 |  |  | 1 | 1 | 0 | 0 |  |
| 2000 | 470 | 790 |  | 1 | 0 | 0 | 0 | 2 | 0 |
| 2001 |  | 841 |  | 0 | 0 | 2 | 0 |  | 1 |
| 2002 | 319 | 286 | 0 | -1 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 59 | 460 | 0 | 0 | 0 | 1 | 0 |  | 0 |
| 2004 |  | 758 |  | 0 | 0 | 0 | 0 | 0 |  |

Tagging data for the 1990 year class

|  |  |  | Recovered |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $\begin{gathered} 2000 \\ \text { release } \end{gathered}$ | $\begin{array}{\|c\|c} 1998 \\ \text { release } \end{array}$ | $\begin{array}{\|c\|} 1997 \\ \text { release } \end{array}$ | $\begin{gathered} 1996 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1995 \\ \text { release } \end{gathered}$ | $\begin{array}{c\|} 1994 \\ \text { release } \end{array}$ | $\begin{gathered} 1993 \\ \text { release } \end{gathered}$ |
| 1994 | 10784 |  |  |  |  |  |  | 0 |  |
| 1995 | 3868 |  |  |  |  |  | 0 | 3 |  |
| 1996 | 6171 | 9009 |  |  |  | 3 | 3 | 9 |  |
| 1997 | 4057 | 9830 |  |  | 2 | 3 | 3 | 7 |  |
| 1998 | 2381 | 2828 |  | 2 | 3 | 1 | 1 | 1 |  |
| 1999 |  | 3402 |  | 3 | 1 | 2 | 2 | 1 |  |
| 2000 | 1219 | 3146 |  |  | 0 | 2 | 2 | 0 |  |
| 2001 |  | 1052 |  | 0 | 0 | 0 | 0 | 2 |  |
| 2002 | 1605 | 1348 | 0 | 1 | 0 | 1 | 0 | 0 |  |
| 2003 | 56 | 1129 | 0 | 0 | 0 | 1 | 0 | 1 |  |
| 2004 |  | 1176 | 0 | 0 | 1 | 1 | 1 | 0 |  |

Table 3.2.4.1 continued
Tagging data for the 1991 year class

| $\stackrel{\curvearrowright}{\approx}$ | $\begin{aligned} & \text { Z } \\ & \text { Z } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { Z } \\ & \text { Z } \\ & 0 \end{aligned}$ | Recovered |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { D} \\ & \text { ర } \\ & \stackrel{0}{C} \end{aligned}$ |  |  | $\begin{gathered} 2000 \\ \text { release } \end{gathered}$ | $\begin{array}{c\|} 1998 \\ \text { release } \end{array}$ | $\begin{array}{\|c\|} 1997 \\ \text { release } \end{array}$ | $\begin{array}{c\|} 1996 \\ \text { release } \end{array}$ | $\begin{gathered} 1995 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1994 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1993 \\ \text { release } \end{gathered}$ |
| 1995 | 21528 |  |  |  |  |  |  |  |  |
| 1996 | 25683 |  |  |  |  |  |  |  |  |
| 1997 | 7129 | 30952 |  |  |  |  | 21 |  |  |
| 1998 | 6002 | 12459 |  |  |  | 6 | 8 |  |  |
| 1999 |  | 14968 |  |  | 4 | 14 | 7 |  |  |
| 2000 | 3802 | 18461 |  | 9 | 1 | 10 | 7 |  |  |
| 2001 |  | 10032 |  | 1 | 2 | 5 | 3 |  |  |
| 2002 | 5878 | 8937 | 10 | 9 | 1 | 1 | 1 |  |  |
| 2003 | 1243 | 9522 | 4 | 7 | 3 | 7 | 4 |  |  |
| 2004 |  | 14288 | 1 | 1 | 4 | 6 | 1 |  |  |

Tagging data for the 1992 year class

| $\underset{\sim}{\aleph}$ | $\begin{aligned} & \text { Z } \\ & \hline \end{aligned}$ | $Z$ $Z$ 0 0 | Recovered |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $\begin{gathered} 2000 \\ \text { release } \end{gathered}$ |  | $\begin{gathered} 1998 \\ \text { release } \end{gathered}$ |  | $\begin{aligned} & 1997 \\ & \text { release } \end{aligned}$ |  | $\begin{gathered} 1996 \\ \text { release } \end{gathered}$ |
| 1996 | 8417 |  |  |  |  |  |  |  |  |
| 1997 | 8353 |  |  |  |  |  |  |  |  |
| 1998 | 22320 | 20695 |  |  |  |  |  |  | 7 |
| 1999 |  | 23790 |  |  |  | 7 |  | 9 | 4 |
| 2000 | 16798 | 31430 |  |  |  | 20 |  | 7 | 15 |
| 2001 |  | 14668 |  |  |  | 8 |  | 0 | 4 |
| 2002 | 9995 | 17305 |  | 12 |  | 23 |  | 2 | 1 |
| 2003 | 2829 | 27306 |  | 11 |  | 11 |  | 4 | 9 |
| 2004 |  | 28022 |  | 19 |  | 17 |  | 2 | 7 |

Tagging data for the 1993 year class

| $\underset{\sim}{\aleph}$ | Z Z B |  | Recovered |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \stackrel{\rightharpoonup}{0} \\ \text { 何 } \\ \text { ad } \end{gathered}$ |  | $\begin{gathered} 2000 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1998 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1997 \\ \text { release } \end{gathered}$ |  |
| 1997 | $\begin{array}{r} 976 \\ 2015 \end{array}$ |  |  |  |  |  |
| 1998 |  |  |  |  |  |  |
| 1999 |  | 8046 |  |  |  | 0 |
| 2000 | 2673 | 9049 |  |  |  | 0 |
| 2001 |  | 3994 |  |  |  | 0 |
| 2002 | 2832 | 5577 | 4 |  |  | 5 |
| 2003 | 1020 | $\begin{aligned} & 6612 \\ & 7315 \\ & \hline \end{aligned}$ | 11 |  |  | 1 |
| 2004 |  |  |  |  |  | 2 |

Table 3.2.4.1 continued
Tagging data for the 1994 year class

| $\underset{\sim}{\widetilde{q}}$ | $\begin{aligned} & \text { Z } \\ & \\ & \hline \end{aligned}$ |  | Recovered |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $\begin{gathered} 2000 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1998 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1997 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1996 \\ \text { release } \end{gathered}$ |
| $\begin{aligned} & 1998 \\ & 1999 \\ & 2000 \\ & 2001 \\ & 2002 \\ & \hline \end{aligned}$ | $\begin{aligned} & 3752 \\ & 2278 \\ & 1143 \\ & \hline \end{aligned}$ | $\begin{aligned} & 2450 \\ & 1104 \\ & 1588 \end{aligned}$ |  |  |  |  |
| $\begin{aligned} & \hline 2003 \\ & 2004 \end{aligned}$ | 442 | $\begin{aligned} & 2154 \\ & 1933 \\ & \hline \end{aligned}$ |  |  | 3 |  |

Tagging data for the 1995 year class


## Table 3.7.2.1.1. Summary of SeaStar exploratory runs - Norwegian Spring-spawning herring..

|  | Default | EstimateM | IncMPlus | TrendLarv | NoLarv | Log | 2002Bar | NoTags | S2And5NoLNoT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spawning stock | 6.632 | 5.495 | 6.575 | 6.091 | 5.873 | 6.642 | 6.642 | 6.905 | 4.739 |
| Juvenile stock | 4.657 | 4.188 | 4.593 | 4.304 | 4.231 | 4.729 | 4.729 | 4.775 | 3.820 |
| terminalF83 | 0.192 | 0.210 | 0.193 | 0.203 | 0.202 | 0.213 |  | 0.163 | 0.222 |
| terminalF90 | 0.315 | 0.434 | 0.316 | 0.347 | 0.354 | 0.598 |  | 0.321 | 0.437 |
| terminalf91 | 0.304 | 0.382 | 0.304 | 0.331 | 0.335 | 0.324 |  | 0.289 | 0.465 |
| terminalF92 | 0.323 | 0.413 | 0.323 | 0.353 | 0.357 | 0.321 |  | 0.332 | 0.442 |
| terminalF93 | 0.448 | 0.522 | 0.448 | 0.472 | 0.478 | 0.531 |  | 0.366 | 0.682 |
| terminalF96 | 0.280 | 0.348 | 0.280 | 0.299 | 0.304 | 0.300 |  | 0.268 | 0.391 |
| terminalF97 | 0.265 | 0.317 | 0.265 | 0.278 | 0.282 | 0.266 |  | 0.255 | 0.359 |
| terminalF98 | 0.137 | 0.165 | 0.136 | 0.149 | 0.154 | 0.137 |  | 0.132 | 0.193 |
| terminalF99 | 0.104 | 0.127 | 0.104 | 0.113 | 0.118 | 0.102 |  | 0.101 | 0.138 |
| terminalF02 | 0.001 | 0.002 | 0.001 | 0.001 | 0.001 | 0.001 |  | 0.001 | 0.002 |
| terminalF03 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| terminalF04 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| cat1 | 1.042 | 1.255 | 1.047 | 1.083 | 1.102 | 1.010 |  | 1.003 |  |
| cat2 | 0.820 | 0.990 | 0.820 | 0.852 | 0.858 | 0.744 |  | 0.797 | 0.958 |
| cat 3 | 0.930 | 1.111 | 0.931 | 0.946 | 0.947 | 0.917 |  | 0.907 |  |
| cat 5 | 1.113 | 1.353 | 1.113 | 1.164 | 1.176 | 1.045 |  | 1.071 | 1.374 |
| catLarvae2 | 4.632 | 5.332 | 4.785 | 5.108 |  | 4.705 |  | 4.533 |  |
| cat 4 | 0.343 | 0.354 | 0.339 | 0.366 | 0.361 | 0.352 | 0.352 | 0.335 |  |
| cat 6 | 0.486 | 0.536 | 0.486 | 0.529 | 0.534 | 0.488 | 0.488 | 0.472 |  |
| catZero | 0.003 | 0.002 | 0.003 | 0.003 | 0.003 | 0.003 | 0.003 | 0.003 | 0.003 |
| distPar1 | 0.485 | 0.477 | 0.486 | 0.481 | 0.485 | 0.518 |  | 0.483 | 0.518 |
| distParLarvae | 0.657 | 0.743 | 0.620 | 0.430 |  | 0.659 |  | 0.652 |  |
| disPar2 | 1.914 | 1.797 | 1.209 | 1.211 | 1.919 | 1.197 | 1.197 | 1.917 | 113.989 |
| distParZero | 0.255 | 0.259 | 0.257 | 0.247 | 0.241 | 0.266 | 0.266 | 0.257 | 0.231 |
| Log-lik pt, S1 | -1.511 | -1.501 | -1.511 | -1.538 | -1.548 | -1.561 |  | -1.479 |  |
| Log-lik pt, S2 | -1.228 | -1.221 | -1.228 | -1.220 | -1.216 | -1.236 |  | -1.249 | -1.157 |
| Log-lik pt, S3 | -1.818 | -1.809 | -1.817 | -1.812 | -1.818 | -1.842 |  | -1.832 |  |
| Log-lik pt, S4 | -3.438 | -3.422 | -3.267 | -3.265 | -3.438 | -3.264 | -3.264 | -3.442 |  |
| Log-lik pt, S5 | -1.526 | -1.490 | -1.527 | -1.521 | -1.517 | -1.554 |  | -1.504 | -1.553 |
| Log-lik pt, S6 | -3.787 | -3.761 | -3.641 | -3.654 | -3.797 | -3.627 | -3.627 | -3.788 |  |
| Num data, S1 | 27 | 27 | 27 | 27 | 27 | 27 | 0 | 27 | 0 |
| Num data, S2 | 58 | 58 | 58 | 58 | 58 | 58 | 0 | 58 | 58 |
| Num data, S3 | 18 | 18 | 18 | 18 | 18 | 18 | 0 | 18 | 0 |
| Num data, S4 | 20 | 20 | 20 | 20 | 20 | 20 | 40 | 20 | 0 |
| Num data, S5 | 65 | 65 | 65 | 65 | 65 | 65 | 0 | 65 | 65 |

## Table 3.7.2.1.1. Summary of SeaStar exploratory runs (con’t)

| Num data, S6 | 11 | 11 | 11 | 11 | 11 | 11 | 22 | 11 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| vpaM | 0.150 | 0.103 | 0.150 | 0.150 | 0.150 | 0.150 | 0.150 | 0.150 | 0.150 |
| vpaMYoung | 0.900 | 1.117 | 0.900 | 0.900 | 0.900 | 0.900 | 0.900 | 0.900 | 0.900 |
| survLogLik1 | -243.917 | -240.733 | -243.981 | -243.723 | -243.662 | -248.002 | -105.169 | -243.099 | -168.020 |
| survLogLik2 | -110.421 | -109.812 | -105.397 | -105.487 | -110.528 | -105.169 | -105.169 | -110.514 | 0.000 |
| tagLogLik1 | -370.238 | -366.024 | -370.199 | -370.085 | -370.088 | -369.623 | 0.000 | 0.000 | -422.105 |
| larvLogLik1 | -79.873 | -83.251 | -78.296 | -68.741 | 0.000 | -79.955 | 0.000 | -79.671 | 0.000 |
| zeroLogLik1 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | -1.236 | 0.000 | 0.000 |
| zeroLogLik2 | -0.704 | -0.352 | -0.772 | -0.278 | 0.065 | -1.236 | -1.236 | -0.798 | 0.606 |

Some abbreviations:
cat<n> Catchability for survey n
Log-lik pt, $S<n>$ Log-likelihood per term, survey $n$
distPar
Distribution parameters (CV for gamma distributions and standard deviation for log normal

Table 3.7.2.1.2. Comparison of number by age in 2004 between the SeaStar assessments in 2005 and 2004. Norwegian Springspawning herring.

| Age in 2004 | Year class | Present <br> assessment | Assessment 2004 |
| :--- | :--- | :--- | :--- |
| 0 | 2004 | 411.694 |  |
| 1 | 2003 | 66.272 | 63.392 |
| 2 | 2002 | 50.903 | 33.527 |
| 3 | 2001 | 0.954 | 1.621 |
| 4 | 2000 | 2.284 | 0.632 |
| 5 | 1999 | 9.568 | 9.601 |
| 6 | 1998 | 9.124 | 10.069 |
| 7 | 1997 | 0.903 | 0.902 |
| 8 | 1996 | 0.714 | 0.964 |
| 9 | 1995 | 0.125 | 0.100 |
| 10 | 1994 | 0.222 | 0.255 |
| 11 | 1993 | 0.548 | 0.600 |
| 12 | 1992 | 1.601 | 2.006 |
| 13 | 1991 | 0.856 | 1.006 |
| 14 | 1990 | 0.109 | 0.210 |
| 15 | 1989 | 0.015 | 0.044 |
| $16+$ | 1988 and older | 0.020 | 0.284 |
|  |  |  |  |

Table 3.7.4.1 Norwegian Spring-spawning herring.. Stock numbers final assessment.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0.000 | 165.952 | 26.803 | 19.801 | 0.797 | 1.847 | 7.476 | 7.661 | 0.596 | 0.464 | 0.077 | 0.129 | 0.301 | 0.998 | 0.543 | 0.069 | 0.019 |
| 2004 | 408.177 | 65.928 | 48.770 | 0.959 | 2.299 | 9.632 | 10.194 | 0.904 | 0.715 | 0.125 | 0.222 | 0.549 | 1.603 | 0.857 | 0.110 | 0.016 | 0.020 |
| 2003 | 162.158 | 119.960 | 2.367 | 2.752 | 11.543 | 12.652 | 1.246 | 1.017 | 0.171 | 0.342 | 0.877 | 2.486 | 1.236 | 0.175 | 0.028 | 0.009 | 0.030 |
| 2002 | 295.054 | 5.821 | 6.866 | 13.625 | 15.393 | 1.723 | 1.533 | 0.231 | 0.498 | 1.304 | 3.604 | 1.802 | 0.260 | 0.046 | 0.018 | 0.001 | 0.051 |
| 2001 | 14.317 | 16.887 | 33.515 | 17.994 | 2.175 | 2.242 | 0.310 | 0.682 | 1.835 | 5.091 | 2.640 | 0.382 | 0.079 | 0.024 | 0.005 | 0.000 | 0.097 |
| 2000 | 41.536 | 82.432 | 44.281 | 2.618 | 3.208 | 0.397 | 0.911 | 2.567 | 7.316 | 4.194 | 0.677 | 0.169 | 0.046 | 0.031 | 0.025 | 0.025 | 0.163 |
| 1999 | 202.751 | 108.915 | 6.446 | 3.876 | 0.500 | 1.204 | 3.446 | 10.229 | 6.127 | 1.101 | 0.310 | 0.069 | 0.079 | 0.037 | 0.124 | 0.000 | 0.355 |
| 1998 | 267.887 | 15.855 | 9.663 | 0.657 | 1.660 | 4.401 | 13.782 | 8.481 | 1.690 | 0.500 | 0.126 | 0.119 | 0.047 | 0.266 | 0.006 | 0.663 | 0.005 |
| 1997 | 38.997 | 23.768 | 1.651 | 2.070 | 5.405 | 17.948 | 12.002 | 2.784 | 0.933 | 0.212 | 0.159 | 0.089 | 0.407 | 0.028 | 1.170 | 0.001 | 0.006 |
| 1996 | 58.460 | 4.060 | 5.138 | 6.316 | 21.622 | 15.638 | 4.249 | 1.522 | 0.357 | 0.191 | 0.112 | 0.544 | 0.051 | 2.261 | 0.002 | 0.000 | 0.011 |
| 1995 | 9.986 | 12.637 | 15.538 | 25.184 | 18.542 | 5.608 | 2.456 | 0.664 | 0.239 | 0.147 | 0.707 | 0.150 | 3.610 | 0.006 | 0.000 | 0.000 | 0.017 |
| 1994 | 31.083 | 38.217 | 61.954 | 21.578 | 6.634 | 3.246 | 0.949 | 0.295 | 0.179 | 0.861 | 0.212 | 4.890 | 0.010 | 0.000 | 0.000 | 0.023 | 0.006 |
| 1993 | 94.008 | 152.383 | 53.085 | 7.738 | 3.887 | 1.197 | 0.352 | 0.212 | 1.033 | 0.267 | 6.123 | 0.012 | 0.000 | 0.000 | 0.027 | 0.000 | 0.010 |
| 1992 | 374.804 | 130.568 | 19.035 | 4.529 | 1.426 | 0.414 | 0.248 | 1.213 | 0.316 | 7.357 | 0.017 | 0.002 | 0.000 | 0.033 | 0.000 | 0.000 | 0.016 |
| 1991 | 321.146 | 46.823 | 11.145 | 1.666 | 0.484 | 0.289 | 1.425 | 0.377 | 8.784 | 0.022 | 0.003 | 0.000 | 0.039 | 0.000 | 0.000 | 0.024 | 0.002 |
| 1990 | 115.167 | 27.413 | 4.122 | 0.582 | 0.339 | 1.668 | 0.449 | 10.449 | 0.027 | 0.005 | 0.003 | 0.048 | 0.000 | 0.000 | 0.029 | 0.001 | 0.002 |
| 1989 | 67.436 | 10.143 | 1.472 | 0.397 | 1.942 | 0.528 | 12.490 | 0.035 | 0.006 | 0.004 | 0.059 | 0.002 | 0.001 | 0.034 | 0.002 | 0.000 | 0.003 |
| 1988 | 24.971 | 3.624 | 0.991 | 2.324 | 0.641 | 15.104 | 0.051 | 0.011 | 0.011 | 0.085 | 0.012 | 0.005 | 0.043 | 0.005 | 0.002 | 0.002 | 0.003 |
| 1987 | 8.935 | 2.447 | 5.772 | 0.766 | 18.089 | 0.079 | 0.017 | 0.021 | 0.128 | 0.027 | 0.016 | 0.055 | 0.014 | 0.009 | 0.010 | 0.001 | 0.004 |
| 1986 | 6.041 | 14.200 | 1.888 | 21.598 | 0.111 | 0.035 | 0.041 | 0.262 | 0.113 | 0.063 | 0.147 | 0.037 | 0.082 | 0.097 | 0.002 | 0.000 | 0.007 |
| 1985 | 34.970 | 4.665 | 53.449 | 0.152 | 0.058 | 0.065 | 0.445 | 0.194 | 0.133 | 0.238 | 0.054 | 0.128 | 0.167 | 0.002 | 0.000 | 0.000 | 0.014 |
| 1984 | 11.528 | 131.465 | 0.378 | 0.072 | 0.081 | 0.583 | 0.246 | 0.168 | 0.294 | 0.071 | 0.167 | 0.201 | 0.002 | 0.000 | 0.000 | 0.026 | 0.000 |
| 1983 | 323.552 | 0.938 | 0.179 | 0.098 | 0.700 | 0.296 | 0.202 | 0.348 | 0.084 | 0.198 | 0.242 | 0.003 | 0.000 | 0.000 | 0.032 | 0.000 | 0.000 |
| 1982 | 2.343 | 0.442 | 0.241 | 0.829 | 0.352 | 0.240 | 0.412 | 0.099 | 0.236 | 0.287 | 0.003 | 0.000 | 0.000 | 0.037 | 0.000 | 0.000 | 0.000 |
| 1981 | 1.100 | 0.595 | 2.057 | 0.413 | 0.283 | 0.487 | 0.118 | 0.279 | 0.343 | 0.004 | 0.000 | 0.000 | 0.044 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1980 | 1.474 | 5.060 | 1.017 | 0.336 | 0.573 | 0.139 | 0.333 | 0.415 | 0.005 | 0.000 | 0.000 | 0.054 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1979 | 12.498 | 2.508 | 0.830 | 0.672 | 0.164 | 0.394 | 0.494 | 0.007 | 0.000 | 0.000 | 0.065 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 |
| 1978 | 6.201 | 2.044 | 1.655 | 0.194 | 0.471 | 0.596 | 0.009 | 0.000 | 0.001 | 0.081 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 |
| 1977 | 5.095 | 4.080 | 0.482 | 0.571 | 0.718 | 0.010 | 0.000 | 0.002 | 0.106 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.003 |
| 1976 | 10.068 | 1.188 | 1.407 | 0.860 | 0.018 | 0.000 | 0.002 | 0.137 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.005 |
| 1975 | 2.971 | 3.467 | 2.117 | 0.024 | 0.000 | 0.004 | 0.192 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.008 |
| 1974 | 8.631 | 5.220 | 0.066 | 0.000 | 0.004 | 0.250 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.014 |
| 1973 | 12.884 | 0.168 | 0.005 | 0.008 | 0.317 | 0.001 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.023 |
| 1972 | 0.957 | 0.077 | 0.051 | 0.407 | 0.005 | 0.006 | 0.003 | 0.001 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.037 |
| 1971 | 0.236 | 0.193 | 1.134 | 0.008 | 0.008 | 0.005 | 0.002 | 0.003 | 0.002 | 0.000 | 0.006 | 0.008 | 0.007 | 0.000 | 0.000 | 0.000 | 0.061 |
| 1970 | 0.661 | 3.620 | 0.072 | 0.017 | 0.026 | 0.003 | 0.007 | 0.005 | 0.002 | 0.021 | 0.038 | 0.038 | 0.000 | 0.000 | 0.000 | 0.000 | 0.101 |
| 1969 | 9.785 | 0.972 | 0.263 | 0.233 | 0.004 | 0.018 | 0.011 | 0.003 | 0.037 | 0.080 | 0.083 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.168 |
| 1968 | 5.187 | 1.333 | 1.182 | 0.111 | 2.048 | 1.509 | 0.018 | 0.144 | 0.238 | 0.469 | 0.003 | 0.002 | 0.002 | 0.003 | 0.004 | 0.003 | 0.276 |
| 1967 | 3.947 | 18.398 | 0.384 | 3.880 | 5.260 | 0.050 | 0.622 | 1.496 | 2.400 | 0.013 | 0.008 | 0.006 | 0.013 | 0.014 | 0.022 | 0.020 | 0.455 |
| 1966 | 51.409 | 1.984 | 12.175 | 8.320 | 0.087 | 1.225 | 3.146 | 5.897 | 0.056 | 0.025 | 0.025 | 0.043 | 0.028 | 0.100 | 0.100 | 0.148 | 0.719 |
| 1965 | 8.491 | 35.874 | 24.938 | 0.198 | 1.700 | 4.271 | 9.222 | 0.086 | 0.045 | 0.037 | 0.071 | 0.076 | 0.224 | 0.233 | 0.321 | 1.627 | 0.148 |
| 1964 | 93.903 | 65.617 | 0.833 | 2.098 | 5.392 | 12.920 | 0.115 | 0.054 | 0.047 | 0.109 | 0.120 | 0.364 | 0.359 | 0.538 | 2.724 | 0.180 | 0.295 |
| 1963 | 168.931 | 5.371 | 8.368 | 7.085 | 15.912 | 0.139 | 0.064 | 0.058 | 0.146 | 0.149 | 0.539 | 0.517 | 0.813 | 4.160 | 0.294 | 0.286 | 0.573 |
| 1962 | 19.003 | 26.983 | 19.058 | 20.474 | 0.170 | 0.078 | 0.075 | 0.192 | 0.186 | 0.690 | 0.658 | 1.070 | 5.710 | 0.390 | 0.391 | 0.517 | 0.589 |
| 1961 | 76.103 | 72.088 | 54.883 | 0.231 | 0.099 | 0.092 | 0.239 | 0.237 | 0.868 | 0.817 | 1.390 | 7.419 | 0.506 | 0.503 | 0.669 | 0.233 | 0.849 |
| 1960 | 197.514 | 156.290 | 1.185 | 0.247 | 0.127 | 0.308 | 0.302 | 1.112 | 1.029 | 1.835 | 9.874 | 0.680 | 0.725 | 0.942 | 0.331 | 0.392 | 1.057 |

Table 3.7.4.1 Norwegian Spring-spawning herring. Stock numbers final assessment (con't)

| 1959 | 412.478 | 3.225 | 1.117 | 0.163 | 0.387 | 0.379 | 1.449 | 1.319 | 2.391 | 12.661 | 0.886 | 0.976 | 1.308 | 0.480 | 0.539 | 0.860 | 1.040 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1958 | 23.094 | 7.136 | 1.447 | 0.468 | 0.460 | 1.804 | 1.628 | 2.988 | 15.760 | 1.105 | 1.266 | 1.737 | 0.665 | 0.709 | 1.075 | 0.948 | 0.959 |
| 1957 | 25.397 | 8.719 | 1.496 | 0.559 | 2.498 | 2.058 | 3.718 | 20.450 | 1.362 | 1.608 | 2.214 | 0.867 | 0.955 | 1.410 | 1.244 | 0.381 | 1.440 |
| 1956 | 29.858 | 6.852 | 2.358 | 3.028 | 2.662 | 4.658 | 26.514 | 1.701 | 2.089 | 2.857 | 1.149 | 1.323 | 1.932 | 1.621 | 0.510 | 0.787 | 1.905 |
| 1955 | 24.971 | 10.303 | 8.247 | 3.193 | 5.710 | 33.009 | 2.099 | 2.631 | 3.616 | 1.427 | 1.745 | 2.564 | 2.102 | 0.656 | 1.005 | 1.124 | 2.629 |
| 1954 | 42.086 | 31.374 | 9.195 | 6.921 | 39.898 | 2.593 | 3.311 | 4.730 | 1.795 | 2.243 | 3.453 | 2.939 | 0.858 | 1.277 | 1.449 | 1.115 | 3.831 |
| 1953 | 86.102 | 30.543 | 17.934 | 47.153 | 3.063 | 3.956 | 5.878 | 2.174 | 2.726 | 4.351 | 3.840 | 1.063 | 1.581 | 1.785 | 1.402 | 3.082 | 4.103 |
| 1952 | 96.644 | 58.460 | 117.910 | 3.601 | 4.661 | 7.479 | 2.673 | 3.387 | 5.465 | 4.869 | 1.321 | 1.930 | 2.190 | 1.744 | 3.782 | 3.938 | 3.488 |
| 1951 | 146.355 | 301.944 | 9.484 | 5.422 | 9.103 | 3.292 | 4.113 | 6.905 | 6.306 | 1.617 | 2.331 | 2.656 | 2.142 | 4.667 | 4.950 | 0.757 | 5.388 |
| 1950 | 750.680 | 26.465 | 14.278 | 10.874 | 4.023 | 4.978 | 8.612 | 8.004 | 1.965 | 2.804 | 3.203 | 2.583 | 5.632 | 6.148 | 0.952 | 2.567 | 6.709 |

Table 3.7.4.2 Norwegian Spring-spawning herring. Summary table SeaStar. Biomass in million tonnes.


## Table 3.9.1.1 Norwegian Spring-spawning herring. Input to the short term prediction



Table 3.9.1.2 Norwegian spring spawning herring. Short term prediction and two optional runs.
Short term:

| SSB | Fbar | Landings |  |  |
| :--- | :--- | :--- | :--- | :--- | Biomass 3+ SSB

Optional 1. Short term, selection on age 4 as on age 5
SSB Fbar Landings Biomass 3+ SSB

| 2006 | 2006 |  |  | 2007 |
| :--- | :--- | :--- | :--- | :--- |
| 2007 |  |  |  |  |
| 6.505 | 0.115 | 0.687 | 11.941 | 8.099 |
| 6.503 | 0.120 | 0.715 | 11.912 | 8.069 |
| 6.500 | 0.125 | 0.744 | 11.883 | 8.039 |
| 6.497 | 0.130 | 0.771 | 11.855 | 8.010 |
| 6.494 | 0.135 | 0.799 | 11.827 | 7.980 |
| 6.491 | 0.140 | 0.827 | 11.799 | 7.951 |
| 6.488 | 0.145 | 0.855 | 11.770 | 7.922 |
| 6.486 | 0.150 | 0.882 | 11.743 | 7.893 |
| 6.483 | 0.155 | 0.909 | 11.715 | 7.864 |
| 6.480 | 0.160 | 0.937 | 11.687 | 7.836 |

Optional 2. Catches in 2006 and SSB in 2007 when a regression of catch to historic Fbar was used.

Fbar Catch SSB
20062007
$\begin{array}{lll}0.115 & 0.741 & 8.027\end{array}$
$\begin{array}{lll}0.120 & 0.772 & 7.990\end{array}$
$\begin{array}{lll}0.125 & 0.803 & 7.954\end{array}$
$\begin{array}{lll}0.130 & 0.834 & 7.917\end{array}$
$\begin{array}{lll}0.135 & 0.865 & 7.881\end{array}$
$\begin{array}{lll}0.140 & 0.895 & 7.845\end{array}$
$\begin{array}{lll}0.145 & 0.926 & 7.810\end{array}$
$\begin{array}{lll}0.150 & 0.957 & 7.774\end{array}$


Figure 3.3.1. Total reported of Norwegian spring-spawning herring in 2004 by ICES rectangle. Grading of the symbols: black dots less than 300 t , open squares $300-3000 \mathrm{t}$, and black squares > 3000 t .


Quarter 1


Quarter 3


Quarter 2


Quarter 4

Figure 3.3.2. Total catches of Norwegian spring-spawning herring in 2004 by quarter and ICES rectangle. Grading of the symbols: black dots less than 300 t , open squares $\mathbf{3 0 0 - 3} 000 \mathrm{t}$, and black squares $>3000$ t. Some catches reported in the northern part of the North Sea in first quarter are considered to be of autumn-spawning origin.


Figure 3.7.2.1.1. Spawning stock history (million tonnes) of Norwegian spring spawning herring as estimated by SeaStar.


Figure 3.7.2.1.2. Estimated year classes and fit to the December survey in Vestfjorden, SeaStar assessment.


Figure 3.7.2.1.3. Estimated year classes and fit to the spawning stock survey along the Norwegian coast, SeaStar assessment.


Figure 3.7.2.1.4. Estimated year classes and fit to the January survey in Vestfjorden, SeaStar assessment.


Figure 3.7.2.1.5. Estimated year classes and fit to the May survey on feeding grounds in the Norwegian Sea, SeaStar assessment.


Figure 3.7.2.1.6. Estimated year classes and fit to the September survey in the Barents Sea, SeaStar assessment.


Figure 3.7.2.1.7. Estimated year classes and fit to the May survey in the Barents Sea, SeaStar assessment.


Figure 3.7.2.1.8. SSB for Norwegian spring spawning herring estimated by SeaStar and fit to larval data. 2003 point below X-axis marks deletion.


Figure 3.7.2.1.9. Quantile-quantile plot for survey and larval data for the SeaStar assessment of Norwegian spring spawning herring.


Figure 3.7.2.1.10. Quantile-quantile plot (blue points) for tag return data for the SeaStar assessment of Norwegian spring spawning herring. The black line shows the simulated theoretical curve for a perfect fit.


Figure 3.7.2.1.11. Quantile-quantile plot (blue points) for tag return data for the SeaStar assessment of Norwegian spring spawning herring when the log normal distribution is assumed for surveys in the Norwegian Sea and along the Norwegian coast.


Figure 3.7.2.1.12. Retrospective plot for the SeaStar assessment of Norwegian spring spawning herring.


Figure 3.7.2.1.13. Retrospective trend in the estimation of catchabilities of acoustic surveys in the SeaStar assessment of Norwegian spring spawning herring.


0 - catch-at-age
1- spawning grounds acoustic in Febr.-March 2-acoust. surv. in wint. area Nov.-December
3- acust. in wintering areas, January
4- Young herring in the Barents Sea (June)
5- Feeding areas, May
6-Young herring in the Barents Sea, September survey

Figure 3.7.2.2.1. Profiles of components of ISVPA loss function for different sources of data (SPALY run).


0 - catch-at-age
1- spawning grounds acoustic in Febr.-March 2-acoust. surv. in wint. area Nov.-December 3- acust. in wintering areas, January
4- Young herring in the Barents Sea (June)
5- Feeding areas, May
6-Young herring in the Barents Sea, September survey
(in thel ast year report there was a missspecification in pictire 6 - it was MDN!!!).

Fig. 3.7.2.2.2 Profiles of components of the ISVPA loss function for "masked" surveys data (masking - from SeaStar)


Figure 3.7.2.2.3 Influence of the outliers' exclusion (using "X-84" rule) on the respective components of the ISVPA loss function.


Figure. 3.7.2.2.4 . Results of NSS herring stock assessment by ISVPA


Figure. 3.7.2.2.5. NSS herring. ISVPA. Comparison to previous assessment




Figure 3.7.2.2.6 NSS herring, ISVPA, retrospective runs.



Figure 3.7.2.2.7. NSS herring. ISVPA residuals. Red ball corresponds to 3




Figure 3.7.2.2.8. NSS herrng. ISVPA. Comparison of results when signal from catch-at-age + surveys 2 and 5 were used to assessments using all data sources


Figure 3.7.2.2.9. Residuals for ISVPA run with 3 sources of information


Figure 3.7.2.2.10. ISVPA. NSS herring. Bootstrap.

Difference of stocknumbers using surveys 2 \& 5: ISVPA - SeaStar (>0 re


Figure 3.7.3.1 Difference of stocknumbers using surveys 2 and 5: ISVPA-SeaStar.


Figure 3.9.1.1 Ratio of catch to SSB regressed to Fbar. The fitted line has an intercept of 0.005 and a steepness of 0.93 .

### 4.1 Stock description

Blue whiting is a pelagic gadoid which is widely distributed in the eastern part of the North Atlantic. The highest concentrations are found along the edge of the continental shelf in areas II, V, VI where it occurs in large schools at depths ranging between 300 and 600 meter but is also present in almost all other management areas between the Barents Sea and the Strait of Gibraltar. The major spawning takes place in February and March, along the edge west of the British Isles. Juveniles are abundant in the Norwegian Sea and Barents Sea. Morphological, physiological, and genetic research has suggested that there may be several components of the stock which mix in the spawning area west of the British Isles (ICES C.M. 2000/ACFM:16;). For assessment purposes blue whiting in these areas is treated as a single stock since it has so far not been possible to define an unambiguous border between populations.

### 4.2 ICES advice and management applicable to 2004 and 2005

In 1998 ICES defined limit and precautionary reference points for this stock: $\mathbf{B}_{\mathrm{lim}}$ ( 1.5 mill.t.), $\mathbf{B}_{\mathrm{pa}}(2.25$ mill.t. $), \mathbf{F}_{\text {lim }}(0.51)$ and $\mathbf{F}_{\mathrm{pa}}(0.32)$. The advice of ICES in following years has been given within a framework defined by these reference points.

In December 2002 EU, Faroe Islands, Iceland, and Norway agreed to implement a long-term management plan for the fisheries of the blue whiting stock, which is consistent with a precautionary approach, aimed at constraining the harvest within safe biological limits and designed to provide for sustainable fisheries and a greater potential yield. The plan should consist of the following:

1. Every effort shall be made to prevent the stock from falling below the minimum level of Spawning Stock Biomass (SSB) of 1500000 tonnes.
2. For 2003 and subsequent years, the Parties agreed to restrict their fishing on the basis of a TAC consistent with a fishing mortality less than 0.32 for appropriate age groups as defined by ICES, unless future scientific advice requires modification of the fishing mortality rate.
3. Should the SSB fall below a reference point of 2250000 tonnes $\left(\boldsymbol{B}_{p a}\right)$ the fishing mortality rate, referred to under paragraph 1, shall be adapted in the light of scientific estimates of the conditions then prevailing. Such adaptation shall ensure a safe and rapid recovery of the SSB to a level in excess of 2250000 tonnes.
4. In order to enhance the potential yield, the Parties shall implement appropriate measures, which will reduce catches of juvenile blue whiting.
5. The Parties shall, as appropriate, review and revise these management measures and strategies on the basis of any new advice provided by ICES

This management plan as a whole has never been implemented. In the absence of agreements on a TAC for 2002, 20032004 and 2005, the Coastal States and the Russian Federation each unilateral implemented catch limits for these years. The combined total of the catch limits greatly exceed the provisions of the agreed management plans and has not lead to a restriction of a further expansion of the fishery on blue whiting. ICES has not evaluated the management plan in relation to the precautionary approach.

In addition to the combined catch limits, agreements are made between countries which regulate access to each others waters. Also technical measures may apply to national fleets or national zones.

In 2003, ICES stated that both estimates of SSB and fishing mortality were high but uncertain. Nevertheless, the spawning stock biomass in 2003 was likely to be above $\mathbf{B}_{\mathrm{pa}}$. Therefore,
based on the most recent estimates of fishing mortality and SSB, ICES classified the stock as likely to be harvested outside safe biological limits ( $\mathrm{F}>\mathbf{F}_{\text {lim }}$ ). The incoming year classes seemed to be strong. ICES recommended that catches should be less than 925000 tonnes in 2004 in order to achieve a $50 \%$ probability that the fishing mortality in 2004 is less than $\mathbf{F}_{\text {pa }}$ (=0.32). This would also assure a high probability that the spawning stock biomass in 2005 to be above $\mathbf{B}_{\mathrm{pa}}$.

In 2004 ICES concluded from the most recent estimates of fishing mortality and SSB, that the stock had full reproduction capacity, but was harvested unsustainably (cat. 2b). Although the estimates of SSB and fishing mortality were not considered precise, it was certain that SSB was above $\mathbf{B}_{\mathrm{pa}}$ and the estimated fishing mortality well above $\mathbf{F}_{\text {lim }}$. Recruitments in the last decade appeared to be at a much higher level than earlier. The (not implemented) management plan implied catches of less than 1.075 million $t$ in 2005 which is expected to keep fishing mortality less than 0.32 with $50 \%$ probability. This would also assure a high probability that the spawning stock biomass in 2006 to be above $\mathbf{B}_{\mathrm{pa}}$. The management plan point 4 calls for a reduction of the catch of juvenile blue whiting which have not been taken place. ICES recommended that measures be taken to protect juveniles.

### 4.3 Description and development of the fisheries in 2004

Total catches figures in 2004 were provided by members of the WG. They were estimated to be 2.4 million tonnes, about the same as last year. Time series with catches by nations and area are given in Tables 4.3.1-4.3.7. No catches from France were reported to ICES for 2004. The Working Group noted that France reported about 19000 tonnes in 2004 to the EU. The information came to late to be included in the assessment.

The spatial and temporal distribution of the catches of blue whiting in 2004 is given by quarter and ICES rectangles in Figure 4.3.2. The distribution of the catch by ICES rectangles for the whole year is given in Figure 4.3.1. In 2004 the catch provided as catch by rectangle represented approximately $100 \%$ of the total WG catch.

### 4.3.1 Denmark

The Danish directed fishery blue whiting fishery is mainly conducted by trawlers using a minimum mesh size of 40 mm . Blue whiting was also taken as by-catch using trawl with mesh sizes between 16 and 36 mm for Norway pout, however in 2004 this fishery was very limited. The directed fishery caught 89500 t mainly in Divisions VIa, IIa, IVa, and Vb. Some blue whiting by-catches are also taken during the human consumption consumption herring fishery in the Skagerak.

### 4.3.2 Germany

The main fleet fishing pelagic species is based at Bremerhaven and Rostock, Dutch owned and operating under German flag. It consists of 3 large pelagic freezer trawlers of sizes greater than 90 m up to more than 120 m with horse power between 4200 and 11000 . Crew consists of about 35 to 40 men. They are specially designed for pelagic fisheries. The catch is pumped into large storage tanks filled with cool water to keep the catch fresh until it is processed.

The Blue Whiting fishery is seasonal in the first and second quarter. The fishery starts in divisions VIIb, cand XII in March and continues in April along divisions VIa and b to divisions Vb 1 and 2 in May. Occasionally there is blue whiting by-catch in the herring directed fishery in IIa and b in August. In 2004 the Atlanto-Scandian Herring directed fishery started in Division IIa in May and continued in IIa and b in June until it ended in IIb in July.

### 4.3.3 Faroe Islands

The Faroese fishery for blue whiting (7 combined purse seiners/trawlers and one large factory vessel) started in January 2004 in the south-western part of the Faroese EEZ (ICES Division Vb ) and on Porcupine Bank area west of Ireland (Division VIIc,k and VIb) in the EU zone. In February the fishery continued in the Porcupine area. Later in February and March the fishery continued in international waters Div. VIb and XII outside the EU zone. In April the fishery had moved northwards towards the Faroese zone (northern VIb and southern Vb). In May the fishery continued in the southern part of the Faroese area and on the slope vest off the Faroe plateau, indicating that the fish migrated west of the Faroes on their way northwards. Later in May the fishery had moved northwest towards the Icelandic border and into the Icelandic zone (Divisions Va and Vb ). In June the fishery operated in the Faroese and Icelandic zones ( Va and IIa) and continued in this area in August, also including the northern part of IIa (International waters). The fishery continued on the Faroe-Iceland ridge in September and October, as well as in the northern part of the International waters (IIa). In November till the end of 2004 the catches were taken in Icelandic and Faroese waters (Va, Vb and IIa), moving closer to the Faroes in December. Over $90 \%$ of the catches were taken with pelagic trawl (44 $\mathrm{mm} \#$ in cod end).

The industrial fleet (5 trawlers) operated mainly in Norwegian waters (ICES Division IVa) with catches of blue whiting scattered throughout the year.

### 4.3.4 France

No information was available to the WG. France did not report any catches to ICES for 2004. The WG was informed that France reported 19467 tonnes to EU in 2004.

### 4.3.5 Iceland

Iceland and Faroes have a bilateral agreement of mutual fishing rights for blue whiting within each other's EEZs. Iceland set a total blue whiting catch quota of 428000 tonnes in 2004 for Icelandic-Faroese and International waters.

The Icelandic directed fishery started in the third week of March in International waters west of the British Isles and a few days later some catches were also taken at southeast Iceland. In April only very few catches were taken in International waters at the beginning of the month and the fishery moved totally into Faroese waters. In May the fishery gradually moved north and by the middle of the monthhad completely shifted into the Icelandic area. In June-October the fishery was almost entirely conducted in the Icelandic zone. The total Icelandic catch was 422079 tonnes compared to 501000 tonnes in 2003. Almost all catches are landed for fish meal.

About 20 purse-seiners/trawlers participated in the Icelandic fishery, using large pelagic trawls with a 40 mm mesh size in the cod end. The length range of the vessels was $51-86 \mathrm{~m}$ with a mean length of 66 m . The engine power range of the fleet was $1943-5520 \mathrm{~kW}$ (2640-7500 HP) with a mean of 3593 kW ( 4882 HP ).

Iceland has set size limitations on landings of blue whiting. If the catch consists of $30 \%$ or more fish smaller than 25 cm , a temporary area closure is imposed.

### 4.3.6 Ireland

The Irish fishery for blue whiting is prosecuted by the pelagic RSW (refrigerated seawater) fleet. Involvement in the fishery began with exploratory fishings in the late 1980s and intensified in the following decade in response to restrictive quotas for mackerel and herring.

Landings peaked in 2004 with around 76,000 t landed from divisions, VIa,b, VIIb, XII and VIIj,k.

Fishing begins in February on the spawning aggregations on the shelf edge west of Ireland and develops northwards towards the Faeroes until the end of the season in April. In 2005 landings were made from 17 trawlers ranging from 40 m to 70 m in length and using trawls up to 1700 m circumferences with a brailler mesh of $35-40 \mathrm{~mm}$. The majority of the catch is landed for reduction to meal but an increasingly important proportion is now processed for human consumption. This is facilitated by the proximity of the fishing grounds to the major Irish pelagic port and processing centre at Killybegs.

### 4.3.7 Netherlands

The Dutch fleet fishing for pelagic species in European waters consists of 14 freezer trawlers and one pair trawler. In addition, a number of flag vessels are operating from the Netherlands. Target species of this fleet are: herring, blue whiting, mackerel, horse mackerel and argentines. Some of these trawlers are also fishing for sardinella and horse mackerel in west African waters during part of the year. The fishery for blue whiting is carried out with large pelagic trawls and is a directed fishery with almost no bycatch of other species. Catches increased in 2004 compared to 2003. Most of the catches in 2004 originated from ICES Division VIa and VIIc and were taken in the first half of the year. All catches are landed frozen for human consumption.

### 4.3.8 Norway

Norway did not set a blue whiting quota for the directed fisheries in the Norwegian EEZ, Jan Mayen zone and international waters for 2004. Likewise, there was no quota regulation for the mixed industrial fishery in the Norwegian EEZ in the North and Norwegian Seas (areas east of $4^{\circ} \mathrm{W}$ ). Through international agreements, 120000 t in the EEZ of EU (of which up to 40000 t could be taken by the mixed industrial fishery in the ICES area IVa) and 36200 t in the Faroese zone were made available to the Norwegian fleet.

The main Norwegian fishery is a directed pelagic trawl fishery, regulated by vessel quotas (in 2004, these were only in effect in the EU and Faroese zone), and is carried out on and west of the spawning areas west of the British Isles and in the Norwegian Sea using pelagic trawls with minimum mesh size of 35 mm . A total of 46 large combined purse seiners/trawlers took part in the fishery in 2004. Blue whiting were also fished in the North Sea and in the southern Norwegian Sea (areas east of $4^{\circ} \mathrm{W}$ ) in the mixed industrial fishery targeting blue whiting and Norway pout. Notice that before 2004 these vessels were only allowed to fish blue whiting south of $64^{\circ} \mathrm{N}$. These vessels use small-meshed trawls operated close to the bottom (minimum mesh size 16 mm ) or pelagic trawls with minimum mesh size of 35 mm .

In 2004, as usual, there was a seasonal progression of the fishery from the international waters off Porcupine Bank and Rockall in the beginning of the season (January-March) towards the shelf edge in EU zone and the banks in the Faroese waters in the end of the spawning season. The fishery in EEZ of EU was stopped 26 April and that in the Faroese zone 16 June after the quotas in the respective zones were taken. The fishery in the Norwegian EEZ, Jan Mayen zone and international waters was stopped for the period 29 April-23 May; this was at least partly a reaction to large proportions of juvenile blue whiting in the catches in the southern Norwegian Sea. For the rest of the season, no regulation took place, although significant proportions of young blue whiting, particularly of age 1 year, occurred in the catches. There were also reports of non-negligible numbers of saithe and redfish being caught as by-catch, but no attempts to estimate the actual quantities.

According to the official statistics, the catch in the directed fishery was 828000 tonnes, which is a new record. In the industrial mixed fishery, 129000 tonnes were taken - a new record as well. Notice than in area IVa, where 106000 t was landed as blue whiting, sampling indicates that about 16000 t represents other species. In contrast to the most previous years, this correction was not reported to the Working Group.

### 4.3.9 Portugal

In 2004 Portuguese landings of blue whiting amounted to 3937 tonnes, increasing by around $50 \%$. Most of the landings were taken by a trawl fishery ( $82.9 \%$ ) The remaining landings were taken by polyvalent ( $16.9 \%$ ) and purse seine ( $0.2 \%$ ) fisheries.

The two main harbours (Portimão and Matosinhos), with tradition of blue whiting landings, have reach $49.2 \%$ of total landings. This last year two other main harbour had importance for blue whiting landings, they were: Sines (in southwest coast) and Vila Real de Santo António (in South coast) with $17.9 \%$ and $16.8 \%$ respectively

### 4.3.10 Russia

The directed blue whiting fishery started at the beginning of the year in the Faroese fishing zone. The fishery in international waters west of Porcupine Bank began on the $8^{\text {th }}$ of February and moved gradually northwards following the spawning migrations. Primo March the fleet left Division VIIc and moved to the Rockall area. (Division VIb) where a successful fishery took place. This fishery ended earlier than in 2003 on $15^{\text {th }}$ of April. This early conclusion of the fishery was due to a more easterly distribution of spawning fish than observed in 2003. Only few catches were taken west of $18^{\circ} \mathrm{W}$ in Division XII where catch rates were somewhat lower than in 2003. Nevertheless, the total catch increased. The number of vessels that participated in this fishery was higher than in 2003, at times the number of vessels were as high as 35. In April the pelagic trawlers returned to the Faroese EEZ (Division Vb 1 ). In summer the vessels moved to the Norwegian EEZs and to international waters (Divisions Vb1 and IIa). From October to December a Russian fleet operated mostly in international waters and in the Faroese fishing zone (IIa, Vb1).

Some blue whiting were caught at Rockall by bottom trawls in the haddock fishery in MarchSeptember. Blue whiting on Rockall Bank was mostly taken as by-catch, predominantly in July and August.

### 4.3.11 Spain

The Spanish blue whiting fishery was carried out mainly by bottom pair trawlers in a directed fishery and by single bottom trawlers in a by-catch fishery, and small quantities ( 10 t ) were also caught by long-liners. These coastal fisheries have trip durations of 1 or 2 days and catches are for human consumption. Thus, coastal landings are rather stable due mainly to market forces.

Pair Bottom Trawl Fishery: The Pair bottom trawl is a traditional fleet that fish mainly blue whiting (above $80 \%$ ) and other pelagic species in Div. VIIIc and North IXa. In the middle of 90 's, VHVO gear (with 25 m of vertical opening) was gradually substituting the traditional one. From 2001 the cod end mesh size was increased to 55 mm . This fleet is composed of 68 pairs ( 136 vessels) with an average of 25 m length, 472 HP , and 145 GRT. The pair trawler fleet landed 13840 t , taken mainly on the border between Divisions VIIIc and IXa.

Bottom Trawl Mixed Fishery: This metier operates in Divisions VIIIc and IXa North, using a cod end mesh size of 65 mm and a vertical opening of 1.2-1.5 m . It targets a wide range of species including horse mackerel, blue whiting, mackerel, hake, anglerfish, megrims, and

Nephrops (the first three species contributes around $70 \%$ of the landings). At present it is composed of around 235 vessels, with an average of 27 m length, 561 HP , and 177 GRT. Bycatches of blue whiting in this fishery were ( 1762 t ).

Spanish landings increased around $13 \%$ in 2004 after the low value observed in 2003, partly explained by the closure of the fishery in some fishing grounds and ports, due to the sinking of the oil tanker "Prestige" in Galician waters in December 2002. These closures took place in the first semester of 2003.

### 4.3.12 Sweden

The total Swedish catch was 19 000t. About $95 \%$ of the catches were caught in the directed Norwegian Sea fishery (ICES Divisions IIa and IIIa) in the 3rd quarter of the year. All the Swedish catches were taken with pelagic trawl, most by single vessels and about 5 000t by pair-trawlers.

### 4.3.13 UK (Scotland)

The total Scottish catch was 58 000t. About $98 \%$ of the catch was caught in the directed spawning fishery (ICES Divisions VIa, Vb, VIb and VIIb) in the first and second quarters of the year. A pelagic fleet of 24 vessels, ranging in length from $45-76 \mathrm{~m}$ with an engine power range of 2149-10720 HP., took part in the fishery. The catch was stored in RSW tanks and mainly landed for fish meal.

### 4.4 Bycatches in the fishery

In last two years, reports were available from Faroes and Iceland on the by-catch of other species, particular cod and saithe, in the pelagic fishery for blue whiting. The knowledge on by-catch is generally very sparse in this fishery, but have recently been investigated by both countries.

### 4.4.1 Icelandic investigations (land based sampling)

During May-December 2004 by-catch in the Icelandic blue whiting fishery in Icelandic and Faroese waters was analysed. From 42 trips ( $10.0 \%$ of all trips) 411 samples were collected in a randomised manner. By-catch species in the samples were quantified and length measured. In general, by-catch was a relatively rare occurrence, but associated with rather wide confidence limits.

The by-catch of saithe and cod were recorded in $35.3 \%$ and $21.7 \%$ of samples, and their total by-catch was $0.69 \%$ ( 2900 tonnes) and $0.26 \%$ ( 1080 tonnes), respectively. by-catch per haul ranging from $0-25 \%$ for saithe and $0-10 \%$ for cod. By-catch of saithe more than doubled since 2003 and that of cod was sevenfold by weight and ninefold as a proportion of total catch (Pálsson et al. 2004).

Spatial distributions indicate that cod is mainly caught as by-catch in Icelandic waters, while by-catch of saithe is found in both Icelandic and Faroese waters, on the Faroe-Iceland Ridge (Figs 4.4.1.1-2). The length distributions indicate that the by-catch mainly constitutes the catchable component (large fish) of the stocks in question (Fig. 4.4.1.3). For saithe especially high by-catch rates on the Faroe-Iceland Ridge, in May-Jul in Icelandic zone and in Nov-Dec in Faroese zone. By-catch rates of cod resulted in highest rates in June-July in Icelandic waters.

An analysis of by-catch in 2003 has been analysed by Pálsson (2005).

### 4.4.2 Faroese investigations at sea

Faroes initiated screenings on board a large factory trawler (M/T Næraberg) in NovemberDecember 2004 (Lamhauge 2004). Relatively high by-catch rates were observed on the FaroeIceland Ridge off the northwestern edge of the Faroe plateau (Fig. 4.4.2.1). Average by-catch rate of saithe was $3.5 \%$ ( 68.2 t ), ranging from $0-20 \%$ on a by-haul basis. The size of the saithe was large, on average 77 cm and 4.7 kg (Fig. 4.4.2.2).

### 4.4.3 Summary

Results from by-catch screenings of saithe and cod in the blue whiting fishery on the FaroeIceland Ridge so far:

- Faroese screening at sea Nov-Dec 2004: saithe 3.5\% (68.2 t), ranging from 0-20\%. Large saithe on average 77 cm and 4.7 kg feeding on blue whiting.
- Icelandic screenings on land 2004: saithe $0.69 \%$ (2900 t) and $\operatorname{cod} 0.26 \% ~(1080 t)$ in 420.000 t of blue whiting caught. Screening of 42 trips or $10 \%$, ranging from $0-25 \%$ by-catch of saithe per haul.
- Especially high by-catch rates on the Faroe-Iceland Ridge in May-July in the Icelandic zone and in Nov-Dec in the Faroese zone.
- Icelandic screenings on land 2003: saithe $0.32 \%$ (1600 t) and $\operatorname{cod} 0.03 \% ~(160 t)$.

The blue whiting fishery in May in the Faroese area has not been screened, and the likelihood of by-catches of saithe might be high.

In general screenings on land might be considered to yield minimum estimates of true bycatch, as some by-catch will be lost or avoided during pumping, as well as some fish will be taken for food on board. This can be seen from the higher by-catch ratio estimates of saithe from the sea based screenings in November 2004 as compared to the Icelandic land based screenings (fished in the approximately same area), where the by-catch of $3.5 \%$ compared to around $2 \%$ in the Icelandic material (estimate from Iceland based on visual inspection of Fig. 4, Pálsson et al. 2004).

In order to minimise by-catch in the pelagic fishery it could be recommended to use a sorting grid in the mid-section of the trawl. This would prevent large fish, such as saithe and cod, to enter the cod-end. However such a method has not been tested properly yet.

### 4.5 Fishery dependent data

### 4.5.1 Sampling intensity

In total 1774 samples were collected from the fisheries in 2004.181235 fish were measured and 27835 were fish aged. Sampled fish were not evenly distributed throughout the fisheries (see text table below).

| Quarter | Fisheries | Directed | Mixed | Southern | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | No. of samples | 356 | 33 | 148 | 537 |
|  | WG Catch | 754678 | 17900 | 4902 | 777480 |
| 2 | No. of samples | 234 | 70 | 159 | 463 |
|  | WG Catch | 835121 | 40450 | 4520 | 880092 |
| 3 | No. of samples | 211 | 55 | 152 | 418 |
|  | WG Catch | 422096 | 45763 | 4368 | 472227 |
| 4 | No. of samples | 147 | 51 | 158 | 356 |
|  | WG Catch | 216646 | 25365 | 5759 | 247770 |
| Total No. of samples |  | 948 | 209 | 617 | 1774 |
| Total WG Catch |  | 2228542 | 129478 | 19549 | 2377569 |

Considering the proportion between catches and sampling, the most intensive sampling took place in the southern fishery of Spain and Portugal. Here one sample was taken for every 32 tonnes, followed by the mixed fishery with one sample for every 620 tonnes, and lastly the directed fishery where there was one sample for every 2351 tonnes caught. In this context it should be noted that implementation of the EU Collection of Fisheries Data, Fisheries Regulation 1639/2001, requires a minimum of one sample to be taken for every 1000 t landed. Detailed information on the number of samples, number of fish measured, and number of fish aged by country and quarter is given in Table 4.5.1.1 and 4.5.1.2. As can be seen, no sampling was carried out by Germany, Sweden and France.

The WG requires the samples to estimate catch in numbers and mean length and mean weight. Therefore, the WG urges all countries that exploit this stock to develop appropriate sampling schemes.

### 4.5.2 Landings

Most of the catches are taken in the directed pelagic trawl fishery in the spawning and postspawning areas (Divisions Vb, VIa,b, and VIIb,c). Catches are also taken in the directed and mixed fishery in Subarea IV and Division IIIa, and in the pelagic trawl fishery in the Subareas I and II, in Divisions Va, and XIVa,b. These fisheries in the northern areas have taken 340 000-2 300000 t per year in the last decade, while catches in the southern areas (Subarea VIII, IX, Divisions VIId,e and g-k) have been stable in the range of $20000-85000 \mathrm{t}$. In Division IXa blue whiting is mainly taken as by-catch in mixed trawl fisheries.

In the last few year the proportion of landings originating from the Norwegian Sea has increased from $5 \%$ in the mid-1990's to $40 \%$ in 2003 and 2004 (Figure 4.5.2.1). This has implications for the stock assessment as much larger proportions of juvenile fish occur in catches from the Norwegian Sea, thus probably changing the exploitation pattern of the fishery as whole.

### 4.5.3 Discards

Discarding of blue whiting is thought to be small. Most of the blue whiting is caught in directed fisheries for reduction purposes. However, some discarding occurs in the fisheries for human consumption and as bycatch in fisheries directed to other species. Reports on discarding and bycatches in fisheries which catch blue whiting were available from Iceland and the Netherlands for the year 2004. Also information on discarding of blue whiting was available from Spain. It was reported to the WG that no discarding of blue whiting occurs in Icelandic, Russian, Faroese and Danish fisheries. No data were available from Ireland and Norway and for all other countries no information was available.

The Netherlands reported discarding of blue whiting in 2004 ( $<4 \%$ in weight) from observations carried out in the EU data collection framework. Most of these discards were reported to be associated with net damage. An annual estimate of discarding of the Dutch fleet is included in the assessment.

Also information of discards was available for some Spanish fleets. Blue whiting is a bycatch in several bottom trawl fisheries directed to a mixture of species. In general the catch rates of blue whiting in these fisheries are low and most of the catch is discarded and only last day catch may be retained for marketing fresh. The estimates of discard in these fisheries in 2004 ranged between 85 and $100 \%$ (In weight). In the directed fishery for blue whiting for human consumption with pair trawls, discards were estimate to be $35 \%$ (in weight) in 2004. No attempt was made to calculate an annual estimate of discarding because of insufficient information of the fleets concerned.

### 4.5.4 Age and length composition of catches

Data on the combined length composition of the 2004 commercial catch by quarter of the year from the directed fisheries in the Norwegian Sea and from the stock's main spawning area were provided by the Faroes, Iceland, Ireland, the Netherlands, Norway, Russia and Scotland. Length composition of blue whiting varied from 9 to 49 cm , with $70 \%$ of fish ranging from $23-28 \mathrm{~cm}$ in length and $98 \%$ from $18-33 \mathrm{~cm}$. The mean length in this fishery was 25.9 cm (Table 4.5.4.1). Length compositions of the blue whiting catch and by-catch from "other fisheries" in the Norwegian Sea and the North Sea and Skagerrak were presented by Norway (Table 4.5.4.2). The catches of blue whiting from the mixed industrial fisheries consisted of fish with lengths of $8-43 \mathrm{~cm}$ and a mean of 14.6 cm . Norway, the Netherlands, Spain and Portugal caught blue whiting in the Southern area. The Spanish and Portugese data used for length distribution of catches showed a length range from $15-40 \mathrm{~cm}$ with a mean length of 21.9 cm (Table 4.5.4.3).

For the directed fisheries in the northern area in 2004, age compositions were provided by Denmark, the Faroe Islands, Iceland, Ireland, Norway, the Netherlands, Russia and Scotland and the sampled catch accounted for $95 \%$ of the total catch. Estimates of catch in numbers for unsampled catches were raised according to the knowledge of how, where, and when the catches were taken. The age compositions in the directed fisheries are given in Table 4.5.4.4.

Age compositions for blue whiting by-catches from "other fisheries" in the North Sea and Skagerrak were provided by Norway and Denmark and sampled catch accounted for $94 \%$ of catches. These data were used for allocation of the remaining part of the total in that area. The age compositions are given in Table 4.5.4.5.

For the fisheries in the Southern area, age composition representing $100 \%$ of the catch were presented by Spain and Portugal. The age compositions in the southern fishery data are given in Table 4.5.4.6.

The combined age composition for the directed fisheries in the Northern area, i.e. the spawning area and the Norwegian Sea, as well as for the by-catch of blue whiting in "other fisheries" and for landings in the Southern area, were assumed to represent the overall age composition of the total landings for the blue whiting stock. The catch numbers-at-age used in the stock assessment are given in Table 4.5.4.7. The 2000 and 2001 year classes were the most numerous in the catches, followed by the 2002, year class. To calculate the total international catch-at-age, and to document how it was done, the program SALLOC was used (ICES 1998/ACFM:18). The allocations are shown in the Annex II.

### 4.5.5 Weight at age

Mean weight-at-age data were available from Denmark, the Faroes, Iceland, Ireland, the Netherlands, Norway, Portugal, Russia, Scotland and Spain. Mean weight-at-age for other countries was based on the allocations shown in the Annex II ("ALLOC" files) and was estimated by the SALLOC program for the total international catch. Table 4.5.5.1 shows the mean weight-at-age for the total catch during 1981-2004 used in the stock assessment. There is a general trend towards lower weight-at-age, except for age-group 10+ years in 2004 that is likely an error. The weight-at-age for the stock was assumed to be the same as the weight-atage for the catch.

### 4.5.6 Length at age

Mean length-at-age data were available from Denmark, the Faroes, Iceland, Ireland, the Netherlands, Norway, Portugal, Russia, Scotland and Spain. Mean length-at-age for other countries was based on the allocations shown in the Annex II ("ALLOC" files) and was
estimated by the SALLOC program for the total international catch. Table 4.5.6.1 shows the mean length-at-age for the 4 main fisheries and the total catch in 2004.

### 4.5.7 Maturity at age

Maturity-at-age used in the assessment was obtained by combining maturity ogives from the southern and northern areas, weighted by catch in numbers-at-age (ICES 1995/Assess:7). These are the same as those used since 1994 (Table 4.5.7.1). Although the values of maturity-at-age probably are too low, sufficient information for estimating new ogives is not available.

### 4.5.8 Natural Mortality

The possible need for revising the current estimate of instantaneous natural mortality rate $M$ for blue whiting was discussed in detail by the 2002 WG. Although it was admitted that the current estimate $M=0.2 \mathrm{yr}^{-1}$ might be too low, the factual basis for revision was ambiguous. More recent methodological work by WGMG (ICES 2003/D:03) emphasises that natural mortality rate cannot be estimated reliably with information normally available for stock assessment models. WG therefore considers that there is no new information that would justify a revision of the current estimate of $M$.

### 4.5.9 CPUE data

### 4.5.9.1 Spanish pair trawl CPUE

The Spanish pair trawls CPUE series (Table 4.5.9.1.1) has been used for several years as a tuning fleet in the blue whiting assessment. This fleet represents only a small part of the landings caught in a small part of the distribution area. Due to this fact, and following a recommendation of the Methods working group (ICES 2003/D:03), this tuning series is not used in the assessments anymore. This year there is not new data about this series so last year's data are presented. Data show a slight decreasing trend in CPUE (Figure 4.5.9.1.1).

### 4.5.9.2 Norwegian CPUE

CPUE data in the spawning area has been collected from the Norwegian commercial fleet 1982-2003; the time series is shown in Figure 6.4.3.2 in the WG report from 2004. No data from 2004 are available to update this time series. The data are not considered to be respresentative for the development of the stock and are not used in the assessment.

### 4.5.10 Effort data

No effort data are available as a measure of activity of the fisheries directed towards the blue whiting stock. The absense of quantitative data demonstrating changes in the fishery is considered to be a handicap in interpreting differences in the results of different assessments.

### 4.6 Fishery independent data

### 4.6.1 Survey abundance indices

### 4.6.1.1 International Blue Whiting survey on the spawning grounds

An international blue whiting spawning stock acoustic survey is carried out on the spawning grounds west of the British Isles in March-April. The survey started in 2004 and is an extension of the Norwegian survey described in section 4.6.1.2. The survey is carried out by Norway, Russia, EU (Ireland and Netherlands), and in 2005 the Faroes joined the survey. This
international survey with broad international participation allowed for broad spatial coverage of the distribution of the stock as well as a relatively dense net of trawl and hydrographical stations. A description of the survey is given in the reference. The survey is coordinated by PGNAPES.

The highest concentrations were recorded in the area between the Hebrides, Rockall and Bill Bailey/Faroes Banks. In comparison to 2004, the bulk of the biomass was observed further offshore in relation to the Hebrides shelf break. The distribution of acoustic backscattering densities for blue whiting as recorded by the six vessels are shown in Figure 4.6.1.1.1

The blue whiting spawning stock estimates based on the international survey are given in the table below. The 2005 estimate of SSB is about $30 \%$ lower than in 2004. An agedisaggregated total stock estimate is presented in Table 4.6.1.1.1, showing that the SSB was dominated by blue whiting of 4 and 5 year olds (year class 2000 and 2001) which contributed about $60 \%$ to the SSB. As in last year, the contribution of juveniles to the stock estimate is low; the survey may not give a good signal of expected recruitment.

| International <br> Survey | Abundance, $10^{9}$ <br> individuals <br> total |  | Biomass, mill. <br> tonnes | Mean <br> seight, | Mean <br> length, |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 137 | 128 | 11.4 | 10.9 | 85.5 | 26.4 |
| 2004 | 90 | 83 | 8.0 | 7.6 | 88.6 | 26.3 |
| 2005 |  |  | total | spawning | g | cm |

Results from the individual vessels are reported in more detail in respective survey reports (O'Donnell et al., WD; Heino et al. b; Ybema; Jacobsen et al.). Of these only the Norwegian time series is long enough to be used in tuning the assessment. For this reason, it was decided to keep the spatial coverage of this survey similar to earlier years. The estimated total biomass and SSB of the international survey and Norwegian time series in 2004 and 2005 are very close.

### 4.6.1.2 Norwegian surveys on the spawning grounds

Norwegian survey on the spawning grounds of blue whiting, west of the British Isles, provides the longest time series covering a significant part of the blue whiting stock, and thereby an important time series for tuning the assessment. While from 2004 onwards this survey has been run as a part of the PGNAPES coordinated international survey (see 4.6.1.1 above), this survey has been run such that the integrity of the existing time series is maintained. The results from both the international and Norwegian survey are very much in agreement.

The estimated total abundance of blue whiting for the 2005 Norwegian survey was 8.5 million tonnes, representing an abundance of $95 \times 10^{9}$ individuals (Table 4.6.1.2.1). The stock estimate obtained in 2005 is significantly smaller than in 2002-2004, both in terms of numbers and biomass (Table 4.6.1.2.2). Virtually all fish were estimated to be mature, and the spawning stock is thus only marginally smaller than the total stock in the area. However, sampling in the areas where juveniles have usually been observed was jeopardized by bad weather, exaggerating the paucity of juveniles in the stock estimate.

A potential caveat in interpreting the results from this survey is the decrease in survey coverage, which in 2005 was about $20 \%$ less than in 2004. To make the numbers more comparable, only those survey strata that were covered in both years are compared and adjusted for relative coverage within each stratum. In 2005, the total stock biomass estimate in those strata that were covered in 2004 is 8.4 million tonnes. Adjusting for slightly smaller coverage within each stratum in 2005 than in 2004, the corrected estimate is 8.7 million tonnes. This estimate has to be compared with the estimate obtained in 2004 for the same
strata, 10.9 million tonnes. The coverage-corrected decrease in total stock biomass is thus about $20 \%$ - a slightly less pessimistic result.

Year class 2000 (age 5 years) continues, albeit with a narrow margin, to be the most abundant year class in the stock, both in terms of biomass and numbers (Fig. 4.6.1.2.1). This was also the dominant year class in 2002-2004, and appears for fourth year in row as the strongest one in record for its age. Abundance of this year class has been reduced by some $37 \%$ from 2004, the year when it presumably was fully recruited to the spawning stock. Year class 2001 is currently almost as abundant as year class 2000, although much less abundant than that year class at the same age. Blue whiting of age 3 years makes also a significant contribution $(\sim 17 \%)$ although it is relatively weak for its age. Year classes 1999 and before lumped together make a similar contribution.

Mean length and weight of blue whiting in the survey area continue to increase (Table 4.6.1.2.2), largely reflecting the increase in the average age and the continued dominance of year class 2000. Condition of young blue whiting was somewhat lower and that of old fish somewhat better compared to 2004; these changes are mostly driven by changes in weight-atage.

### 4.6.1.3 Russian surveys on the spawning grounds

The Russian research vessels surveyed the blue whiting spawning stock since 1980. The area was firstly located between $57^{\circ} \mathrm{N}$ and $69^{\circ} \mathrm{N}$, and the works conducted in April-May. Since 1984 area was shifted to $50^{\circ}-62^{\circ} \mathrm{N}$, and March-April become to be as new period. Mean area covered yearly in 1981-96 amounts 53 thousands miles ${ }^{2}$. The highest abundance of blue whiting was observed in 1986. The vessels operated using pelagic and partly a bottom trawls with cod-end mesh 16 mm based on a distance of 30 -miles between the latitudinal transects. After 1990 the surveys fulfilled in coordination with Norwegian research vessels. The time series used in exploratory runs is shown in Table 4.6.1.3.1.

In 2005 two research vessels operated west of British Isles in spring. "Fridtjof Nansen" carried out an acoustic survey mostly inside of the EEZ zone from $53^{\circ}$ to $61^{\circ} 40^{\prime} \mathrm{N}$ on 18.3-14.4 based on a distance of 30 -miles between the latitudinal transects. "Atlantniro" fulfilled the same work outside of EEZ zone from $54^{\circ} 33^{\prime}$ to $60^{\circ} 07^{\prime} \mathrm{N}$ on $15.3-08.4$ based on 15 -miles distance between the transects. The total blue whiting biomass in the international waters was estimated as 1 million tones. This is similar to the equivalent result in 2002 and almost three times lower than in 2003. Environmental conditions are suggested as reasons for the fluctuations. Mean temperature in the depth layer 400-600 m on Rockall was $9.56^{\circ} \mathrm{C}$ during the survey in 2002 and $9.64^{\circ} \mathrm{C}$ in 2005 while in 2003 the temperature reached $9.76^{\circ}$. In both years the east directed transport of water predominated putting obstacles in the western migration of spawners.

### 4.6.1.4 International ecosystem survey in the Norwegian Sea

The international ecosystem survey in the Nordic Seas is aimed at observing the pelagic ecosystem in the area, with particular focus on herring, blue whiting, mackerel, zooplankton and hydrography. In addition the Norwegian Sea was covered during June-July and in August 2005 on a national basis. The observations on herring and blue whiting are done by acoustic observation with main focus on Norwegian spring-spawning herring (ASH) and blue whiting in the Norwegian Sea. The survey is carried out in May since 1995 by the Faroes, Iceland, Norway, and Russia, and since 1997 (except 2002 and 2003) also the EU. The survey is used in the final assessment for the first time.

In 2005, blue whiting were observed in most of the survey area, with the highest densities off the north-western Norway and in the south, between the Faroes and Norway and the Faroes
and Iceland (Figure 4.6.1.4.1). There is a tendency of mean length to increase away from the Norwegian coast towards northeast. Both distribution and size structure of the stock are broadly similar compared to the survey in previous year.

Stock estimate for the total survey area is given in Table 4.6.1.4.1. Blue whiting of age 1 year dominate the stock both in terms of numbers and biomass. The stock biomass estimate, 6.6 million tones, is $36 \%$ lower than in $2004,10.4$ million tonnes. Also stock numbers are decreased, from $152 \cdot 10^{9}$ in 2004 to $120 \cdot 10^{9}$ in 2005 . These rather dramatic decreases are largely due to the more restricted coverage in the south-western part of the survey area where post-spawning fish aggregate at the time of the survey. For the standard survey area that has been covered each year (between $8^{\circ} \mathrm{W}-20^{\circ} \mathrm{E}$ and north of $63^{\circ} \mathrm{N}$ ) the estimate is 4.7 million tonnes, down $14 \%$ from 5.4 million tonnes measured in 2004 . The stock estimate in numbers at $95 \cdot 10^{9}$ is virtually unchanged from 2004. The proportions of large and old blue whiting are slightly lower in the standard survey area than in the total survey area; this is expected as the post-spawner aggregations in the southwest are largely excluded from the standard area. Time series of stock estimates for the standard area are given in Table 4.6.1.4.2.

### 4.6.1.5 Iceland acoustic summer survey

Iceland has carried out an acoustic survey in summer in the Icelandic EEZ since 1998. No survey was carried out in 2005. Age-disaggregated results are available from 1999 onwards until and including 2004 (Table 4.6.1.5.1). The survey in 2003 gave stock estimate that was well above the one from the previous years. However, this survey had a wider coverage than those before. This reflects wider distribution of blue whiting in the Icelandic waters. It can be argued that the increase reflects genuine increase of blue whiting in the area (see the discussion in Heino et al., WD 2004). Table 4.6.1.5.1 gives nevertheless also an estimate that is calculated for the more restricted area covered all years. The survey in 2004 confirms the results from 2003 that year class 2003 is weak in Icelandic waters. Year class 2004 is average in stregth. The mean age is the highest observed in this survey.

### 4.6.1.6 Icelandic Groundfish survey 2005

The Icelandic Groundfish survey has been conducted in every March month since 1985. The survey covers the Icelandic continental shelf down to 400 m and in the period 1985-1995 and since 2004 the Icelandic part of the Icelandic-Faeroes ridge was covered. The number of stations hys been variable or between 500 and 600 . The survey is conducted by $4-5$ trawlers all identical trawlers built in Japan in the early 1970.

Blue withing is caught in the survey near the edges of continental shelf in the southern part of the survey area. In recent years blue withing has been found at 70-100 in the survey. Blue withing was not measured regularly in the survey until 1996. The length distributions and age readings from the same month indicate that smaller than 21 cm fish can be considered a proxy of age 1 fish. The indices of age 1 fish are calculated as the mean number of age 1 blue withing per station in the areae covered by the survey in all years (Table 4.6.1.6.1). The survey is not used for prediction of recruitment.

### 4.6.1.7 Norwegian sea summer survey

In 1981-2001 Norway ran an acoustic survey in the Norwegian Sea in order to follow the migration of Norwegian spring spawning herring and to measure blue whiting in its feeding areas. The stock estimates in numbers at age are given in Table 4.6.1.7.1. This survey used to give the first indication of the incoming year class measured at age 1; in 2004 SGAMHBW recommended using indices from this survey at ages 1-4 years. However, as the survey is terminated it provides little information for the latest years in the assessment, and it was decided not to use the survey in the final assessment.

### 4.6.1.8 Russian survey in the Barents Sea

A blue whiting stock survey was carried out in the Barents Sea and adjacent waters in October-December 2004. This is the continuation of survey series started in 2000. Two vessels participated in the survey: "Smolensk" and "Fridtjof Nansen". As the work mentioned was a part of multi-species trawl-acoustic survey, both bottom and pelagic trawls were applied to the schools identification. The area investigated makes 236637 square miles. Estimated biomass of blue whiting over the whole surveyed area constituted 524 thousand tones whereas in 2003 only 350 thousand tones were found. Almost whole biomass was distributed in the ICES divisions IIa and IIb (Fig. 4.6.1.8.1). The survey is not used in the assessment.

### 4.6.1.9 Spanish bottom trawl survey

Bottom trawl surveys have been conducted off the Galician (NW Spain) coast since 1980, following a stratified random sampling design and covering depths down to 500 m . The survey directed to a misxture of species. Since 1983, the area covered in the Spanish survey was extended to completely cover Spanish waters in Division VIIIc. A new stratification has been established since 1997. Stratified mean catches and standard errors are shown in Table 4.6.1.9.1 . Larger mean catch rates are observed in the $100-500 \mathrm{~m}$ depth range. Since 1988 the highest catch rates in the Spanish survey were observed in 1999 ( $124 \mathrm{~kg} / \mathrm{haul}$ ). The 2004 estimate is $25 \mathrm{~kg} / \mathrm{haul}$. (Figure 4.6.1.9.1). The survey is not used in the assessments.

### 4.6.1.10 Portuguese bottom trawl survey

Bottom trawl surveys have been conducted off the Portuguese coast since 1979, following a stratified random sampling design and covering depths down to 500 m . The area covered in the Portuguese survey was extended in 1989 to the 750 m contour. Stratified mean catches and standard errors from the Portuguese survey are shown in Table 4.6.1.10.1. Larger mean catch rates are observed in the $100-500 \mathrm{~m}$ depth range. The 2004 estimate is around average in the Portuguese autumn survey ( $84 \mathrm{~kg} /$ haul) . The Portuguese autumn surveys generally give higher values than in the summer surveys, and a better correlation with the Spanish surveys (Figure 4.6.1.10.1). The survey is not used in the assessments

### 4.6.1.11 Faroes plateau spring bottom trawl survey

On the Faroe plateau an annual demersal bottom trawl surveys is carried out during spring (March 1996-2005). The survey is not used in the assessments. The survey is aimed at cod, haddock and saithe, but varying amounts of blue whiting are caught as by-catch each year. The size of the blue whiting ranges between $11-45 \mathrm{~cm}$ during spring, including 1 -group and older fish.

After the spawning west of the British Isles and Ireland in spring the larvae drift northwards with the currents passing the Faroes and young of the year are found in the deeper regions of the shelf in August. One group blue whiting are also found on the shelf in the deeper regions and are caught in both the autumn and spring groundfish surveys.

Unfortunately no otoliths are available from these groundfish surveys prior to 2004, making the separation of the 1 - and 2 -group blue whiting inaccurate. The 1 -group blue whiting were visually separated from the length distribution from the spring surveys (Table 4.6.1.11.1). In the spring surveys the 1 -group is usually well separated in the length distribution (Table 4.6.1.11.2). In 2004 the 1 -group was separated from otolith samples during the survey.

### 4.6.1.12 Faroes plateau autumn bottom trawl survey

On the Faroe plateau an annual demersal bottom trawl survey is carried out in autumn (August-September 1994-2004). The surveys is aimed at cod, haddock and saithe, but varying amounts of blue whiting are caught as by-catch each year. The size of the blue whiting ranges between $9-45 \mathrm{~cm}$ during autumn, including 0 -group and older blue whiting.

Unfortunately no otoliths are available from these groundfish surveys, making the separation of the 0 - and 1 -group blue whiting inaccurate. For the summer survey Icelandic age-length keys (from June-August) were used to split the length into age groups (Table 4.6.1.12.1). Since the Icelandic age-length keys were from the June-August period, they could not be used to split the spring data (see previous section).

The consistency in the time series in representing the 0 - and 1 -group can be checked by a comparison of the abundance of the 0 -group caught in summer with the abundance of 1 -group the following spring. Figure 4.6.1.12.1 shows the bottom trawl 0 -group index (number of blue whiting caught $10^{-3}$ ) from autumn 1996-2004 compared to the corresponding 1-group spring index lagged one year to match the 0 -group index, the correlation is relatively low, the $r^{2}$ is around 0.41 . The survey is not used in the assessments.

### 4.6.1.13 Norwegian shrimp survey in the North Sea

Norway has conducted a bottom trawl survey on shrimp in the northern parts of the North Sea annually in October 1984-2002 (Tveite 2000). Blue whiting is caught in $>95 \%$ of trawl hauls, with individual lengths ranging between 9 and 52 cm . Blue whiting have not been aged, but based on sampling of the commercial catches in the area by Norway, a reasonably good separation of 0-group blue whiting can be achieved by assuming that all individuals less than 19 cm in body length are of age 0 years. Separation of other age groups has not been attempted. The results are summarized in Table 6.4.2.6 of the WG report from 2004, showing that 0 -group blue whiting are occasionally very abundant, but that the years of great abundance show only weak correspondence with large year classes seen in the analytic assessment. This applies in particular to the recent strong year classes that have not occurred in large numbers in the North Sea. The survey is not used in the assessments.

### 4.6.1.14 Norwegian bottom trawl survey in the Barents Sea

Norway has conducted bottom trawl surveys targeting cod and other demersal fish in the Barents Sea since late 1970s. From 1981 onwards there have been systematically designed surveys carried out during the winter months (usually late January-early March) by at least two Norwegian vessels; in some years the survey has been conducted in co-operation with Russia. Blue whiting is a regular by-catch species in these surveys, and is in some years one of the most abundant species. Most of the blue whiting catches (or samples thereof) have been measured for body length, but very few age readings are available (from 2004 onwards otoliths are systematically collected). The existing age readings suggest that virtually all blue whiting less than 20 cm in length belong to 1 -group and that while some 1 -group blue whiting are larger, the resulting underestimation is not very large. An abundance index of all blue whiting and putative 1 -group blue whiting from 1981 onwards is given in Table 4.6.1.14.1 and follows methods described in Heino et al. (2003) except that all blue whiting $<20 \mathrm{~cm}$ in length are taken as 1 -group blue whiting (threshold 21 cm was used in 2004). 1-group indices for both 2004 and 2005 are relatively high, ranking as fifth and third highest in this 25 -year time series, although well below the record from 2001. Also the total indices are at a historically high level.

### 4.6.1.15 Russian Egg survey

Ichthyoplankton samples were collected during the spring cruises of R/V "Atlantniro" and "Fritdjof Nansen" in 2005. A vertical plankton net lowered to 600 meter depth. Altogether more than 70 samples were fixed, 50 of them were treated before the WG meeting. The egg development stages were determined according to scale of Seaton \& Bailey (1971). Stage V predominated in the samples (Table 4.6.1.15.1). Survey started near the Porcupine bank a few days after the main spawning finished there. The number of eggs in the catches in that part of area was not large for the given reasons. The vessels arrived at the next spawning ground located between $55^{\circ}$ and $56^{\circ} 30^{\prime} \mathrm{N}$ after peak spawning activity but a lot of eggs were still available. The greatest catches were derived at Rockall where the survey and intensive spawning were falling at the same time (Figure 4.6.1.15.1). Following the traditional approach (Hensen, 1887; English, 1964) the data on the duration of spawning period and on the incubation time were used to convert egg number to the SSB. Seasonal egg production was calculated by equation:
$\mathrm{P}=\frac{N_{i} t_{i}}{T_{i}(1-y)}$; where
P - seasonal production; $N_{i}$ - number of eggs in the area surveyed; $T_{i}$ - duration of egg development; $y$ - daily loss of eggs.
$\mathrm{T}_{\mathrm{i}}$ assumed as 4.8 days and $\mathrm{t}_{\mathrm{i}}$ as 59 days to obtain comparable results with Coombs (1979) estimation. Daily loss assumed as 5\% (Bailey, 1974). Surfer-8 software applied to count total number of eggs in the volume investigated. Egg production in the area from which the samples are elaborated amounts to $1,8 \times 10^{14}$ eggs. Using the actual data on fecundity (38000), mean weight of fish $(0,99 \mathrm{~kg})$ and sex ratio (1:1) the spawning stock biomass on the area investigated was estimated as $1,8 \times 10^{14} \times 0.099 \times 2 / 38000=0.94$ mln.tones. That number is corrected for a relation between the biomasses inside and outside of area surveyed according to results of international acoustic survey in 2005. The area from which the samples were treated contains 1.8 of the total 8 mln . tones. Hence, total SSB have to be 4.2 mln . tonnes.

The main uncertainties in this calculation are related with mortality rate and development time of eggs. Cautious levels of those parameters are used above. Present-day temperatures should determine more fast incubation. Mortality had to grow as well (Coombs, 1977). Daily loss rates are usually especially high on the 3-rd stage of development. One attempt to avoid those gaps was undertaken by use the 1-st and 2-nd stages only. In this case development time may be assumed as 1 day, and the impact of mortality ignored. $N_{i}$ may be suggested as equal to the one at the most numerous stages. Those is a stage V. Fish at stage V were some days ago at the stages I and II while their number were at least same as at the stage V. If samples were collected several days earlier, this number would be correspondingly more. However in this case the spawning duration should be estimated in different way. Instead of the whole period when the spawning indications are available, the time of active spawning should be used when large schools of spawning fish are observed - 45 days. The mentioned corrections are included into the estimation. By given data the SSB is counted as 5.6 mln . tones.

The approach used is stochastic as is the case with other ways of stock assessment. Increasing the number of samples and improving the coincidence of yhe survey with peak spawning could improve results. Still despite the rare net of tows and patched egg distribution, the preliminary figures obtained are not far from results of acoustic and analytical estimates. Therefore the egg surveys continuation seems to be a reasonable. The survey resuls cannot be used to tune the assessment at present.

### 4.7 Stock Assessment

The catch at age data were explored using a number of assessment tools. Apart from catch curve analyses, exploratory assessments were carried out using AMCI, ICA, ISVPA and SMS.

### 4.7.1 Evaluation of data underlying the assessment

## Catches

Figure 4.7.1.1 show the number of blue withing caught plotted on log scale and curves corresponding to $\mathrm{Z}=0.6$ drawn for comparison. The picture seems to indicate that yearclasses 1995 and 1996 have been disappearing from the catches at a rate corresponding to Z higher than 0.6 (around 0.7 ) but yearclass 1996 at a rate around 0.7 . This interpretation of the graphs is dependent on the fishing effort being reasonably constant in recent years but gradually increasing fishing effort makes the slopes lower than the real value of Z . The catch curves also indicated that the smaller year classes may be exploited at a slightly lower rate. The curves are on the other hand difficult to interpret for agegroups that have not been fully recruited to the fishery for more than 2 years, which applies to most of the agegroups that are relevant today.

Figure 4.7.1.2 shows $\log$ of the abundance indices from the Norwegian survey on the spawning grounds in February. According to the figures the yearclasses 1998 and 1999 have been disappearing at a rate corresponding to $\mathrm{Z}=0.6$, but yearclass 1996 at higher rate of approximately $\mathrm{Z}=0.7$. For younger year classes interpretation of the curves is difficult.

## Surveys

Acoustic survey on the spawning grounds in February.
Figure Figure 4.7.1.2 shows the internal consistency of the survey by comparing the abundance indices of a yearclass in the survey to the abundance indices of the same yearclass the year before. The consistency is best between age 3 and 4 but detoriates after that, most likely due to increasing fishing mortality which is variable through the period. For the oldest ages sampling in the survey is most likely a problem. Only a small part of each yearclass is mature at age 2. This part is variable from year to year leading to variability in the indices for age 2 which are therefore not as good indicators of stock size as the abundance indices of older fish. Figure 4.7.1.3 shows the abundance indices from the spawning survey plotted on $\log$ scale. The figure indicates that in recent year $\mathrm{Z}>0.6$ for the oldest age groups.

Survey in the Norwegian sea in May.
Figure 4.7.1.2 indicates that the use of indices for age 2 from the spawning fish acoustic survey might be questionable and the survey does not at all cover age 1. The May acoustic survey in the Norwegian sea and adjacent areas is covering large part of the immature part of the stock and might therefore give better indices of recruitment. The survey is on the other hand not considered as an reliable indicator of the mature part of the stock which migrates much more than the immature part.

Figure 4.7.1.4 shows that the internal consistency of the May survey is reasonably good but there are some exceptions like for yearclass 2001 between ages 1 and 2 and the 1996 yearclass between ages 4 and 5. The figures as tables from the survey do also indicate that the abundance indices decrease rapidly with age.

### 4.7.2 Data Exploration with AMCI

Explorations were carried out with the latest version of AMCI, version 2.4. In comparison to version 2.3a that was used in 2004, the new version contains some new functionality (not
used) and bug fixes. Repeating the final AMCI assessment from 2004 with the new version of AMCI yielded essentially identical results.

The final WG assessment from 2004 updated with the new data available (catch in 2004 and Norwegian spawning stock survey in 2005) - so-called SPALY run - shows an overall increase in SSB (and a corresponding decrease in F) since about year 2000 in comparison to the 2004 assessment (Figure 4.7.2.1). Recruitment estimates from about year 1999 onwards are also slightly increased. However, the overall patterns remain very similar.

In tuning the assessment, three time series are traditionally used: Norwegian and Russian acoustic spawning stock surveys and the Norwegian Sea summer acoustic survey. However, Russian spawning stock and Norwegian Sea summer surveys were terminated respectively in 1996 and 2001. Furthermore, Norwegian spawning stock survey is split in two because of changes in instrumentation (the same is true for the other two surveys). It has been asserted that the surveys that do not extend to the most recent years contribute very little information to the assessment and can be dropped. Indeed, AMCI run with the modern part of the Norwegian spawning stock survey (1991-2005) as the only tuning time series yields results that are very similar to the results from the SPALY run. All further AMCI explorations were therefore tuned without the terminated time series.

Two recruitment surveys were tested: international acoustic survey in the Norwegian Sea in May, and winter bottom trawl survey in the Barents Sea. Age range 1-2 years from the international May survey was used, while at the same time the lower age for the Norwegian spawning stock survey was increased by 1 year; the age range then is $3-8$ years. With these changes, SSB in the most recent years is somewhat increased, and the recruitment in 2001 is somewhat higher. However, observed and modelled yield do not fit well in 2002-2003.

It was also tried to use the Barents Sea data. However, because of the strong non-linearity between modelled recruitment (final AMCI run in 2004 WG ) and 1 -group index from the Barents Sea, the data cannot be directly included in AMCI. A two-tier procedure was then employed: a linear model on $\log -\log$ scale was used to estimate the power linking these two time series in 1994-2002 (in earlier years, larger meshes were used), and then using this relationship to construct a time series of modelled recruitments from the Barents Sea data that best fit the 2004 assessment. This time series was then used as an additional tuning time series. The resulting assessment shows slightly lower SSB and recruitment than the corresponding run without the Barents Sea data. However, the residuals from the fitted values do not add up to zero, raising some concerns about lack of convergence or other technical problems.

To increase the fit between observed and modelled yield in 2002-2003, the weight of the yield partial objective was increase from 1 to 2 . This was presented as the final AMCI run and is shown in Figure 4.7.6.1. Model residuals are shown in Figure 4.7.2.2. Retrospective analysis is illustrated in Figure 4.7.2.3 and the results from bootstrapping catches and survey time series in Figure 4.7.2.4 and Figure 4.7.2.5. Furthermore, the results turned out to be rather robust with respect to changes in relative weighting of different ages in the catch and survey data.

The residuals show no worrisome features, and the retrospective analysis suggests that the assessment is not prone to bias. As always, there is increasing uncertainty towards the end of the assessment period. Bootstrapping suggests that a range of combinations of $F$ in the final year and SSB in the end of the final year ranging from low F and high SSB to high F and low SSB are possible given the uncertainty in the data. Among other things, the uncertainty in recruitment in 2001 (probably the strongest year class in record) is manifested in the estimate of SSB in the end of the final year.

Judging the assessment against external information reveals some worrisome features. The estimated exploitation pattern (not shown) does not show increase in fishing mortality of young fish (ages 1-2 years) that would be expected from the increase of the fishery in the Norwegian Sea with different selectivity. Also the decrease in SSB from 2004 to 2005 is much less than suggested by the Norwegian and international spawning stock surveys. Also the SSB estimate is very high in comparison to the levels prior to 2001 and not in accordance with the sentiment and purported lower catch rates from the fishery in the spawning grounds in 2005 and in the Norwegian Sea in summer. Recruitment from 2000 onwards is estimated to be considerably higher than the Barents Sea recruitment index suggests.

All in all, the final AMCI assessment is acceptable when judged with conventional model diagnostics, but does not quite match the external information available. It is quite likely that this is related to the model assuming roughly constant exploitation pattern from year to year, an assumption that is likely grossly violated. Although AMCI allows for some changes in exploitation pattern, the changes may have been stronger than the model can deal with. The assessment is therefore to be judged as particularly uncertain.

### 4.7.3 Data Exploration with ISVPA

If to retain the same model settings as were taken in 2004 stock assessment by ISVPA (SPALY run), i.e., effort-controlled version of the model with unbiased separable representation of fishing mortalities and single selection pattern for the whole period, profiles of components of the loss function, corresponding to different sources of information, reveal minima only for catch-at-age and the second period of Norway spawning acoustic (Figure 4.7.3.1), the minimum for the second period of Norwegian Sea acoustic survey, which was present in last year assessment, now has deteriorated. This survey was not undertaken after 2001 and changes in fishing mortality in 2004 cannot now influence sufficiently and clearly (in presence of all kinds of noise) the goodness of fit to these data which are now rather distant from terminal year.

What is also can be seen in Figure 4.7.3.1 is that now minimum for catch-at-age is somewhat shifted with respect to the minimum of survey 2 (in the last year assessment they were closer to each other).

If compare the results of the catch-controlled version with the last year results for this version of the model (Figure 4.7.3.2), it can be seen that positions of minima for catch-at-age and survey 2 in this year assessment are almost coincide, analogously to what was observed in last year assessment. From this point of view the catch-controlled version may looks preferable for this year stock assessment. Signal from survey 6 has deteriorated this year analogously to the results of the effort-controlled version.

But comparison of ISVPA-derived results from stock assessments in 2004 and 2005 years (Figure 4.7.3.3) shows that results of the effort-controlled version are more in line with the previous assessment. Besides, results for this version less dependable on exclusion of survey 6 (compare families of F controlled and catch controlled curves on Figure 4.7.3.3, corresponding to effort- and catch-controlled versions). Bearing in mind also that in the previous assessment the effort-controlled version was shown to be more retrospective stable, it look reasonable not to change the model settings this year, but only to exclude survey 6 , which now gives no signal along with other auxiliary information, except second part of Norwegian spawning stock acoustic surveys.

Also it is interesting to note that the estimates of $\mathrm{R}(1)$ for 2004 are very close to each other for both versions of the model.

Results of stock assessment by the effort-controlled version of the ISVPA, where signals only from catch-at-age and second part of Norwegian spawning stock acoustic surveys are shown,
in Figure 4.7.3.4. The estimated abundance of age group 1 for 2004 ( 21469 millions) is almost equal to 1981-2003 historical arithmetic mean for estimates of $\mathrm{R}(1)$ ( 21569 millions) and is not far from the median (16 153 millions) and the geometric mean value (16 178 millions).

Results of retrospective runs are presented on Figure 4.7.3.5.
Comparison of theoretical catches with reported values for are presented on Figure 4.7.3.6. As in last year assessment, theoretical catches in 2002 are higher than reported ones. In agreement with its ideology, the effort-controlled version does not consider catch-at-age data as true and theoretical analogues of catch-at-age may deviate from reported values.

Bootstrap estimates of uncertainty in the results are presented on Figure 4.7.3.7.
Figure 4.7.3.8.represents residuals in catch-at-age and surveys. Residuals in catch-at-age have cohort effect what indicates that estimates of abundance for some cohorts are less based on reported catch-at-age and more based on theoretical values, derived from estimated selection pattern. If estimation of abundances are more based on theoretical catch-at-age (through estimates selection pattern), cohort effect in catch residuals may outline cohorts with problems in catch-at-age data and shows that overall estimate for cohorts with higher residuals is less based on reported catch-at-age data.

### 4.7.4 Data Exploration with ICA

The data was explored with a number of different scenarios to try and obtain the best fit to the model. Explorations began with a same procedure as last year (SPALY) run (run 1) on updated fishery and survey data. The analysis planned to investigate the effect of:

- The different tuning fleets on model fit and stability
- Different length for the separable period
- Different levels of selection (S)
- Different relative weights of age groups 1 and 2
- Different reference age for separable period

The results of these different runs are presented Figure 4.7.4.1.
In exploring the contribution of the different tuning fleets the WG began by removing all tuning fleet data except the most recent series, i.e., the Norwegian spawning ground acoustic survey. Values for SSB and F appeared unchanged from the SPALY run. There was no indication of increased uncertainty in the run results indicating that inclusion of the old tuning fleet data is unnecessary. This run (run 2) was taken as the departure point for further exploration.

The effect of using a reduced separable period of 3 years was investigated in run three with all other parameters from Run 2 retained. An additional run (run 4) with the same separable period but with altered values for relative weights at age 1 and 2 and an increased value for fixed selection on the last age was conducted. The thinking behind reducing the separable period to three years was to see if better resolution for SSB and F could be achieved during a period when there had been a change in exploitation pattern due to expansion of the fishery into the Norwegian sea in the second quarter. By down weighting the value for relative weights at ages 1 and 2 it was hoped to reduce noise in the results. This noise may be a spawning ground fishery effect where the youngest year classes are not fully recruited. The result of the reduced separable period (run 3) was to raise the estimated SSB substantially and lower the value for F. Noise in the model output increased and the model fit deteriorated. This was indicated by changes in the residual pattern. Compared with a longer separable period of 8 years used in runs 1 and 2 the results of these runs appeared less stable. Down weighting the
relative weights of ages 1 and 2 (run 4) and increasing the $S$ value had no discernible effect on the output.

In Run 5 the separable period was raised to 4 years and the reference age increased from 3 to 5. The value for selection was raised to the maximum. and relative weights for ages 1 and 2 down $n$ weighted to 0.5 . This produced similar values and model fit for SSB and $F$ to runs 1 and 2 though with a slight deterioration in the quality of the diagnostics.

Run 2 showed the best-defined minimum value for SSQ and offered a high level of consistency with the previous years assessment. Taking this as the final ICA run shows that SSB continues at a high level. F is also comparatively high. Recruitment is well below the very high values of recent years though still at or above the best values estimated for the earlier part of the time series.

The stock summary and diagnostics from final run 2 are presented in Figures 4.7.4.2. to 3 . Comparisons of the different ICA runs are presented in Figure 4.7.4.4;

Retrospective analysis (Figure 4.7.4.4) shows that the final runs for ICA have been developing consistency over the last four years. Running the same settings with 2001 and2000 as the final year show a quite different stock characterised by much lower SSB and raised F values.

| Settings used for ICA final run |  |  |
| :--- | :---: | :---: |
| Number of age structured tuning series | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ |
| Number of biomass tuning series | 3, each split | 1 |
| Number of years for separable constraint | 0 | 0 |
| Reference age for separable constraint | 8 | 8 |
| Constant exploitation pattern | 3 | 3 |
| S to be fixed on last age | Yes | Yes |
| Age range for mean F | 1.5 | 1.5 |
| Catchability model for tuning fleets | $3-7$ | $3-7$ |
| Age range for the analysis | Linear | Linear |
| Survey weights for all fleets | $1-10$ | $1-10$ |
| Shrinkage | $100 \%$ | $100 \%$ |
| Manual down weighting | No | No |
| Tuning series split | Yes | Yes |
| Weighting of age 1 catch numbers | Yes | Yes |

### 4.7.5 Data Exploration with SMS

SMS (Stochastic Multi Species model) (Lewy and Vinther, 2004) is an age structured assessment model to handle biological interaction, however, it can be reduced to operate with one species only. In "single species mode" an objective functions for catch at age numbers and survey indices at age time series are minimized assuming a log-normal error distribution for both data sources. SMS uses maximum likelihood to weight the various data sources. For more details see section (XX).

Since the last WG meeting, SMS has been modified slightly. Now, it is possible to give a starting and an ending date for surveys and observed indices are compared with the mean stock number within that period. Previously, the mean stock numbers within the calendar year was used. In addition, it is now possible to use survey indices for the year following the last assessment year. In such cases, it is assumed that the survey is conducted the 1. January in all years. By using this assumption there is no need to guess on the mortalities in the year following the assessment year.'

## SPALY run

A SMS run was made using the same settings as last year and the updated data series.

## Catch data

Catch data for the period 1981-2004, age 1 to age 10+ were used. The age selection pattern was assumed constant within two periods, 1981-1992 and 1993-2004. Selection at age was estimated by individual age group for age 1 to 7 and combined for age 8-10. Variance of catch observation was estimated separately for age 1 , age 2, age 3-6 combined and age 7-10 combined.

## Tuning data

Survey data and settings used in the last year's SMS run are presented in the table below. Last year, it was not possible to use survey data sampled after the assessment period and for this first run the "Acoustic survey on spawning grounds" used data for the period 1981-2004.

| Survey | Year range | Catchability at age | Variance at age of survey <br> observation |
| :--- | :--- | :--- | :--- |
| Acoustic survey on <br> spawning grounds | $1981-1990$ <br> $1991-2005$ | By age-group: age 2-6. <br> Combined: age 7-8 | By age: age 2, <br> Combined: 3-5 and 6-8 |
| Russian acoustic survey | $1982-1991$ | By age-group: age 3-6. | Combined: 3-5 and 6-8 |
|  | $1992-1996$ | Combined: age 7-8 |  |
| Norwegian Sea | $1981-1990$ | By age-group: age 1-5. | By age: 1-2 and 5 |
|  | $1991-2001$ | Combined: age 6-7 | Combined: 3-4 |

The results of a model run with the same settings as last year and the updated input data are given in Figure 4.7.5.1. Compared with last year's assessment the 2003 SSB is now estimated one million tonnes higher and the 2003 fishing mortality is reduced by $25 \%$.

## SPALY settings with use of the 2005 survey data

SMS is now able to use survey data from the year after the last assessment year. Using this feature and the 2005 data from the acoustic survey on the spawning grounds gives a slightly higher F and lower SSB (Figure 4.7.5.1) for the most recent years compared to the SPALY run. Recruitment is estimated lower for the 2000-2003 year-classes when the 2005 survey data are used, which results in an almost one million tonnes lower SSB in 2005The extension of the assessment time series up to 2004 seems to be the main cause for the upward revision of the 2003 SSB. The additional use of 2005 survey data does not affect the stock estimate up to 2004 much, but the estimate for the SSB in 2005 is reduced by $15 \%$.

SMS estimates relative low CV for catch observations and a rather high CV for survey indices (Table 4.7.5.1). The Russian acoustic and the Norwegian Sea acoustic surveys have in general the highest survey CV and the two surveys do not cover the most recent assessment years. The overall contribution to assessment model from these two surveys is therefore limited.

## SMS using only the spawning grounds acoustic survey.

The results from a SMS run where data from the acoustic survey on the spawning ground were the only survey data are included in Figure 4.7.5.1. The difference from the results using 3 surveys fleets is less than $1-2 \%$ for the main output values and the CV of SSB and estimated stock numbers at age in the terminal year is not affected by more than a fraction of a percent. Moreover the number of model parameters is reduced from 109 to 87. It was therefore decided not to use the Russian and the Norwegian Sea acoustic surveys in the next SMS exploratory runs.

## SMS using spawning ground and juvenile surveys

The acoustic survey conducted in the Norwegian Sea in May, 2000-2005 may give information on the juveniles, even though the time series is short. Data for the 1 and 2-groups were used in a new run.

The effects of including the survey is an increase in the estimate on juveniles in the terminal assessment years and the SSB estimated for 2005 (Figure 4.7.5.1)

## Diagnostics

The SMS diagnostics (Table 4.7.5.2) show that the log-likelihood contributions are highest from catch data. For the surveys the likelihood contributions are highest from the Norwegian Sea May survey and for the most recent survey period of the survey on the spawning grounds. The same information can be found in the estimated CV for catch and survey observations, where the catch observations have the lowest CV. The estimated selection pattern (labelled F, age effect) indicates a shift towards the age group 3-5 in the most recent separable period (Figure 4.7.5.2). Survey catchabilities at age from the spawning area are fairly constant for the younger ages for the two periods, but considerably lower in the most recent period for the older age-groups (Table 4.7.5.2). This confirms the decision to split of the survey into time periods.

There is no consistent pattern in the residuals plot for catch observations (Figure 4.7.5.3) even though the plot indicates a possible shift in the exploitation pattern for the oldest ages in the last four years. Survey residuals (Figure 4.7.5.4) have a very clear year effect, often seen for acoustic time series. The residuals for the 2-group in 2005 are rather large with higher than expected survey CPUE from "juvenile survey" and much lower than expected for the spawning grounds survey. The spawning ground survey might not be a good survey for the 2group, and the CPUE of 2-groups in 2005 should maybe have been seen as an extreme outlier and deleted. Observed and predicted yield are fairly overlapping, however with some discrepancy for the most recent years (Figure 4.7.5.5)

The retrospective pattern for 1998-2004 (Figure 4.7.5.6), excluding the very short "Norwegian Sea, May 2000-2005" survey shows a tendency for overestimation of $F$ and underestimation of SSB. There was no survey data for 1997, which may cause the very unstable estimation of F in the period 1998-2000. After that period the estimated of F and SSB are rather stable from year to year. SMS uses maximum likelihood to estimate the parameters and put as default an equal a priori weighting factor $(=1.0)$ on both catch at age and survey at age information. To investigate the sensitivity of this default settings, different a priori weights were applied to the survey observations ranging from a low weight (weighting factor 0.1) to a high weighting factor (3.0). Weighting on catch observation was kept constant at 1.0. The results (Figure 4.7.5.7) show that the estimated SSB are sensitive to the a priori weighting factors on surveys. That shows that the overall signal in the two data sources, catch and survey, is the same, however the details differs slightly. From Figure 4.7.5.7, it is clear that an increasing a priory weight on survey observations gives an increase in the estimated SSB in the terminal year. Said in another way, survey observations indicate a higher SSB than the catch data.

When the default a priori weighting factors are applied, catch observation get the lowest CV (Table 4.7.5.2) which shows that the maximum likelihood method see catch data as more precise than survey data. Increasing a priori weight on surveys gives increasing CV on catch observation, and a priori weight on surveys higher than 3.0, result in CV on catches higher than $100-150 \%$. This means that there is almost no information in catch data which results in highly variable estimated SSB based purely on surveys. Results for a priori weight on surveys higher than 3.0 are not shown on Figure 4.7.5.7.

## Results

Estimated fishing mortality in Table 4.7.5.3, stock numbers are presented in Table 4.7.5.4, stock summary in Table 4.7.5.5 and Figure 4.7.5.8 and stock recruit plot in Figure 4.7.5.9. The uncertainty of estimated stock numbers in 2004, SSB and F are presented in Figure 4.7.5.10 and 4.7.5.11. A CV at $35 \%$ for the estimated 1 -group stock number and $22 \% \mathrm{CV}$ for the 2group emphasize that rather little is known about the abundance of the recruiting ages to the fishery. However, compared with the estimated CV in last year's SMS assessment the inclusion of the May survey for juveniles has reduced the uncertainties on the juveniles considerably, as the uncertainties last year were estimated to $45 \%$ and $30 \%$. The use of the 2005 survey on the spawning grounds has reduced the CV on SSB in the last assessment years. CV of SSB in 2003-2004 are now $10 \%$ and $12 \%$, a clear reduction from last years assessment with CV at SSB at $14 \%$ and $22 \%$.

The uncertainties presented above have been calculated from the Hessian matrix. Assessment uncertainties estimated by the used of Markov Chain Monte Carlo (MCMC) simulations, with 200000 chains thinned by a factor 500 is presented on Figure 4.7.5.12 and 4.7.5.13. The median of average $F(0.57)$ in 2004 is higher than the $F(0.51)$ estimated from SMS using maximum likelihood.

### 4.7.6 Comparison of results of different assessments

The results of the preferred assessments, carried out with each assessment model (AMCI, ICA, ISVPA and SMS), are compared in Figure 4.7.6.1. The different assessments give basically the same results as last year with the exception of the most recent years. In general the results of the models are in good agreement with each other. All models use the assumption of separability. Differences between the results may to a large extent originate from the differences in external data sources used in the assessment or the capability of the model to use certain kind of data. Also different weighting of the various input sources (example: see Figure 4.7.5.7) may have contributed to differences in the results.

SSB: All models show the same development of the SSB in the historical time series. They indicate a significant increase in SSB in the late 1990's to a historic high in 2003. At its maximum size the stock is estimated by the different models between 5.7 and 6.9 million t . In 2004 and 2005 the SBB declined. The high SSB are in line with the results from the acoustic survey on the spawning grounds although there is a difference in trend between the survey and the assessments in 2001-2003. The decline in SSB in 2005 indicated by the assessment corresponds to the survey. The stock in 2005 is estimated between 4.8 and 5.5 million t .

Fishing mortality: Trends in fishing mortality estimated by the different models are show the same trend but are noisier than SSB estimates. ISVPA estimates of F between 1986 and 1990 are consistently higher than those estimated by the other models. All models indicate a sharp decrease in F from 1990 to 1991. The reasons for this decrease are unclear and may reflect a shift in the fishery to other components of the stock. Also, all models indicate F has increased after 1994. Between 2000 and 2003 trends in F become unclear. In 2004, F is estimated to increase by all assessments and is well above $\mathbf{F}_{\mathrm{pa}}$ and close to $\mathbf{F}_{\text {lim }}$.

Recruitment: All models indicate a significant increase in the level of recruitment after 1995. The 1996 year class was estimated as the strongest in the time series. The estimate of this year class is almost the same by all models. The trends in the time series by ISVPA differ somewhat from the other assessment models. The estimates of recruitment in 2002 and 2003 estimated by ICA are higher than for the other models. The recruitment is 2004 is not estimated well by any model or has been assumed. Recruitment for this year will be estimated separately.

Catches: All models estimate catches in those periods where separability of the fishing pattern has been assumed. Figure 4.7.6.1. shows that there is in general good agreement between observed and modelled catches in the historical period. However, there are differences are in the most recent period. All models estimate a higher catch in 2002 than the reported one and a lower catch than reported in 2003 and 2004. ISVPA also show noticeable discrepancies between observed and modelled catches in some earlier years.

Comparision with last year's assessment: Compared to last year all tuning methods estimate somewhat lower F and higher biomasses in recent years (Figure 4.7.6.2 and 4.7.6.3). SMS and AMCI show the largest differences. The ICA assessment is the most consistent. The change in levels has also been a problem in previous years and contribute to fluctuations in prognoses and the stability of the advice which is based on absolute reference points.

The WG felt it difficult to make a choice between the models. There is little information supporting a specific choice between them. Nevertheless, it was necessary to make a choice as a basis for the predictions. None of the models investigated showed a possible shift in exploitation by the expansion of the fishery into the Norwegian Sea in the most recent 4 years. All models appear to be sensitive for the inclusion of a new year's data and show in some cases large residuals between observed and estimated parameter values. There still appear to be very strong year effects in all surveys and similar in all models whereas ISVPA also show clear cohort effects exist.

The WG selected SMS as the preferred assessment. This choice of the WG differs from last year where AMCI was selected. The choice for SMS was based on the fact that is could demonstrate the sensitivity of the model to the weight given to the survey and the catch data. Also the historical and recent estimates were always close to those estimated by most other models.

### 4.7.7 Final Assessment

The key settings and data for the final blue whiting assessment in 2005 are shown in the table below. The key settings of the final assessment in 2002-2004 are also shown for comparison. Some of the settings are described in more detail after the table.

| Settings/options for the AMCI run | 2002 | 2003 | 2004 | 2005 |
| :---: | :---: | :---: | :---: | :---: |
| Software | AMCI 2.1 | AMCI 2.2 | AMCI 2.3a | SMS |
| Age range for the analysis | 0-10+ | 1-10+ | 1-10+ | 1-10+ |
| Last age a plus-group? | Yes | Yes | Yes | Yes |
| Age at recruitment (from Jan 1 in the year of spawning) | 0.5 | 1 | 1 | 1 |
| Recruitment in the terminal year | Fixed | Fixed | Fixed | Estimated |
| Recruitment in the terminal year-1 | Estimated | Fixed | Estimated | Estimated |
| Catch data |  |  |  |  |
| Weights for the partial objective functions for the catch fleet |  |  |  |  |
| Log sum of squares of catches-at-age | 1 | 1 | 1 | Estimated |
| Log sum of squares of yearly yields | 1 | 1 | 1 | 0 |
| Weights of catch-at-age, age 0 and 1 years | 0.1, 0.5 | n.a., 0.5 | n.a., 0.5 | n.a., estimated |
| Constant selection pattern for the catch fleet? | Almost | Almost | Almost | 2 periods |
| Selectivity for age 10 equals average of selectivity at age 899 ? | No | Yes | Yes | Yes |
| Age-structured tuning time series |  |  |  |  |
| Norwegian acoustic survey on the spawning grounds ages 2-8 | 1981-2002 | 1981-2003 | 1981-2004 | 1981-2005 |
| Flat selectivity for ages 6-8? | Yes | No | No | No, 7-8 |
| Weight in tuning for the partial objective function | 1 | 1 | 1 | Estimated |
| Russian acoustic survey on the spawning grounds ages 3-8, | 1982-1996 | 1982-1996 | 1982-1996 | not used |
| Flat selectivity for ages 7-8? | Yes | No | No | n.a. |
| Weight in tuning for the partial objective function | 1 | 1 | 1 | n.a. |
| Norwegian Sea summer acoustic survey, ages 1-7 | 1981-2001 | 1981-2001 | 1981-2001 | not used |
| Flat selectivity for ages 5-7? | Yes | No | No | n.a. |
| Weight in tuning for the partial objective function | 1 | 1 | 1 | n.a. |
| Norwegian Sea international ecosystem survey, ages 1-2 | not used | not used | not used | 2000-2005 |
| Weight in tuning for the partial objective function | n.a. | n.a. | n.a. | Estimated |
| Biomass tuning time series | 0 | 0 | 0 | 0 |

Survey data used in tuning are shown in Table 4.7.7.1. As in previous years, the Norwegian acoustic survey on the spawning grounds was split into two time periods reflecting a likely change in catchability caused by a change in acoustic equipment (from Simrad EK-400 to EK500). Survey indices are treated as relative abundance indices.

Fishing mortality was modelled as strictly separable within two periods, 1981-1992 and 19932004, in contrast to earlier years when small gradual changes in selection were allowed.

Recruitment in 2004 was estimated. Recruitment in 2005 was set to $43.7 \times 10^{9}$ on the basis of the Barents Sea 1 -group index from winter 2005 calibrated with the SMS-estimated recruitment in 1981-2003 (see Section 4.8). In earlier years, recruitment in the final assessment year has not been estimated but set to a mean value calculated from earlier recruitments (usually 10-year geometric mean).

Catch-at-age data are input at yearly resolution (Table 4.5.4.7). The spawning stock is derived from the stock numbers January 1st (was the first quarter in the 2004 assessment), and the survey indices are related to the mean values in the survey period (Table 4.7.7.1). The yearly fishing mortality is assumed constant over a year (in the 2004 assessment, was split on quarters assuming that the proportion 0.35 of the total annual fishing mortality occurs in the first and in the second quarter, 0.2 in the third quarter, and 0.1 in the fourth quarter).

The model was run until 2005. The SSB in 2005 is a predicted value based on assumed recruitment in $2005\left(43.7 \times 10^{9}\right.$, see above). The key results are presented in Tables 4.7.5.44.7.5.6 and summarized in Figure 4.7.5.9. Residuals of the model fit are shown in Figures 4.7.5.3-4.7.5.4. Some modest cohort effects are visible in the catch residuals for the early cohorts. More importantly, there are age-specific blocks of positive and negative residuals for the latest years (i.e., systematic differences in modelled and observed catches), suggesting
changes in selection pattern. Year effects occur throughout the spawning stock survey time series.

The assessment (Tables 4.7.5.6, Figure 4.7.5.9) indicates that fishing mortality first steadily increased since mid-1990s and has then been fluctuating between $\mathbf{F}_{\mathrm{pa}}$ and $\mathbf{F}_{\text {lim }}$; the most recent value is similar to $\mathbf{F}_{\text {lim }}$. The exploitation pattern (Figure 4.7.5.2) indicates only a modest change between the two periods. SSB has been at a historically high level since 1998 and is only marginally less than the record in 2003. Year class 2000 (recruits in 2001) is the highest in the time series by a large margin. All year classes 1995-2004 are strong in comparison to recruitment before 1995: even the weakest year class born after 1996, that of 1998 , is much stronger than was typical before 1995.

Retrospective analysis (Figure 4.7 .5 .6) shows a small but consistent bias towards underestimation of SSB and over-estimation of fishing mortality. Marko Chain Monte Carlo (MCMC) simulations (Figure 4.7.5.12) give an indication of the uncertainty in the assessment. Temporal patterns in recruitment, spawning stock biomass and fishing mortality are reproduced, but uncertainty in the absolute level of these metrics during the recent years is clearly visible. Nevertheless, the qualitative development of fishing mortality (marked increase) and spawning stock biomass (marked decrease) from 2003 to 2005 appear to be robustly estimated. The $95 \%$ confidence intervals for SSB in 2004-2005 are wide but do not overlap with $\mathbf{B}_{\mathrm{pa}}$, whereas fishing mortality in 2004 is with $>95 \%$ probability larger than $\mathbf{F}_{\mathrm{pa}}$, with modal F close to $\mathbf{F}_{\text {lim }}$. The estimates of terminal F (2004) and SSB (2005) are negatively correlated (Figure 4.7.5.13). The two-topped density distribution of SSB in 2004 indicates uncertainty in assessment.

### 4.8 Recruitment estimates

Information on recruitment (age 1 year) in particular and young blue whiting (ages 0-2 years) in general is provided by a number of surveys, none of these covering the whole distribution of juvenile blue whiting. Only one of these, international Norwegian Sea ecosystem survey in May, is used in the current assessment. The shortcoming of this time series is its shortness, a problem that applies to most other time series as well. Here we discuss this and other surveys from the viewpoint of recruitment in the latest years.

Figure 4.8.1 shows an overview of normalized survey time series in the northern areas (the Faroes and northwards). Information on age 0 is scanty (only two time series) and does not provide a consistent signal. Information on age 1 year is provided by several surveys. These all tend to suggest that recruitment in all years between 2000-2005 has been close to (20022003) or above (2000-2001, 2004-2005) average of the respective time series. At age 2 years the data are scanty but suggest a decreasing trend.

Figure 4.8.2 shows that the internal consistency of the Norwegian Sea ecosystem survey in May is reasonably good but there are some exceptions like for year-class 2001 between ages 1 and 2 years and the 1996 year-class between ages 4 and 5 years. The graphs also indicate that the abundance indices decrease rapidly with age (slope <1).

Another potential problem with the Norwegian Sea ecosystem survey (other than short time series) is that it does not cover the whole nursery area of blue whiting. Acoustic measurements in Icelandic water from July 2000-2004 indicate that part of the nursery areas are near Iceland (section 4.6.1.5). The part that grows up in Icelandic waters is variable but the Icelandic indices of age 1 year are between 10 and $50 \%$ of the indices for the same age group in the Norwegian Sea (highest for year-class 2001). If blue whiting of age group 1 year is considered reasonably stationary in its behaviour, adding the Icelandic July data to the Norwegian Sea May data could be possible. Acoustic measurement of blue whiting in Icelandic waters was not conducted in 2005. The data from the May survey for the area west of $18^{\circ} \mathrm{W}$ were used as
proxy for age 1 but in this survey little 1 -group blue whiting was measured (section 4.6.1.6). For age 2 years the Icelandic survey was not used.

Use of bottom trawl surveys as a measure of recruitment is another possibility that could be investigated. Data on blue whiting are available from the winter bottom trawl survey in the Barents Sea since 1981 (section 4.6.1.14). Figure 4.6 .3 shows those indices from the survey plotted against the recruitment estimate from 2004 WG on both log and ordinary scales. The relationship can be modelled by a power curve (estimated power 2.8) or it can be argued that year-classes smaller than about 30000 million at age 1 year are not properly detected in the survey. Although the Barents Sea is not considered a major nursery area for blue whiting (in summer 2004, estimate at age 1 year was about $5.8 \times 10^{9}$, about $20 \%$ of the estimates in the Norwegian Sea surveys in May but of similar to the estimates in the Icelandic July surveys), the survey could be measuring the component drifting into the Norwegian Sea. This is supported by figure 4.6 .4 that shows the index of age 1 year in the Barents Sea vs. the acoustic abundance of age 1 year in the Norwegian Sea.

The Icelandic groundfish survey in March (section 4.6.1.15) commenced in 1985 but length measurements of blue whiting started in 1996 so indices of age 1 year blue whiting can only be calculated since 1996. In figure 4.6 .4 it can be seen how the indices fit with other indices of recruitment and with the assessment. The immediate answer is that they do not fit well with the assessment nor with the other surveys. They are on the other hand representative for another area and should possibly be added to the other indices as suggested for the Icelandic acoustic measurement. Figure 4.6 .5 shows the result from that kind of model with the estimated model combining Icelandic and Barents Sea bottom trawl surveys as $N_{1}=0.76 I_{i c e}+0.124 I_{b a r}$. The figure indicates a reasonable prediction of recruitment except for year-class 1997 that is missing in both surveys.

The longest time series for predicting recruitment is the Barents Sea winter survey. A possible complication is the change in mesh size from 35-40 to 18 mm in 1994. We make recruitment predictions by fitting model $\log$ (WGrecruitment) $\sim \mathrm{a}_{0}+\mathrm{a}_{1} \log$ (Index) with data from 1994 onwards, or model $\log$ (WGrecruitment) $\sim \mathrm{a}_{0}+\mathrm{a}_{1} \log ($ Index $)+\mathrm{a}_{2}$ Mesh where Mesh is an indicator variable having value 1 in 1981-1993 and 0 later on. As the WG recruitment estimates we use final assessments from 2004 and 2005. These models are illustrated in Figure 4.8.6.

To summarize no good recruitment survey is yet available but a number of surveys can be used to predict the size of year-classes 2003 and 2004 at age 1. The following table shows recruitment estimates for those year-classes using the different surveys. All these are well above the long term average and close to or higher than the average during the recent "high productivity" period.

| Survey data | Calibration <br> data | Time period | Recruitment <br> $2004\left(10^{9}\right)$ | Recruitment <br> $2005\left(10^{9}\right)$ |
| :--- | :---: | :---: | :---: | :---: |
| Barents Sea | WG04 | $1994-2002$ | 33 | 35 |
| Barents Sea | WG05 | $1994-2003$ | 41 | 44 |
| Barents Sea | WG04 | $1981-2002$ | 32 | 34 |
| Barents Sea | WG05 | $1981-2003$ | 40 | 42 |
| Norwegian Sea ecosystem | WG05* | $2000-2005$ | 42 | 48 |
| Norwegian Sea May+Icelandic <br> acoustic | WG05* | $2000-2005$ | 41 | 45 |
| Barents Sea+Icelandic groundfish | WG05* | $1996-2005$ | 32 | 32 |

[^0]
### 4.9 Forecast

### 4.9.1 Short term forecast

Last year short term projection was carried out using the STPR3 (Skagen, 2002), using bootstrap output from the final AMCI assessment as input. This year SMS was used to produce the final assessment, and this program cannot provide bootstrap data. Therefore, the previous STPR3 approach could not be continued. Instead, it was decided to make a deterministic short term forecast based on the SMS maximum likelihood estimate and in addition a stochastic projection based on MCMC functionality build in to SMS (SMS is using the ad-Model builder software, which provides MCMC).

By using the MCMC a number of sets of SMS model parameters are produced, where the correlation between the parameters are maintained. 200 sets were produced from 100000 MCMC chains thinned by a factor of 500 . This parameter set (the posterior distribution) was used as basis for 200 replicate short term projections, which due to the different parameter values, will produce variable initial stock size for the projection and variable exploitation pattern and F status quo. SMS has no model for mean weight in the sea, catch mean weight or proportion mature, and these values were kept constant in the projections. The mean and confidence limits of the predicted SSB, and Yield were calculated from the output of the 200 MCMC projections.

Input
Mean weight at age in the sea, mean weight in the catch, natural mortality and proportion mature were derived from the average of the values for the period 2001-2004. The SMS assessment assumes a constant exploitation pattern for the period 1991-2005, and this was used in the projection. Recruitment at age 1 in 2005 was assumed at $43.7 * 10^{\wedge} 6$ (see Section 4.8). A geometric mean recruitment and its variance in 1996-2003 was used for recruitment in 2006. Table 4.9.1.1 gives an overview of the input data.

For the fishery in 2005 it was assumed that
the catch will reach 2 million tonnes;
the F status quo F will applied

## Ouput

The predicted catch and SSB for two options are presented in Table 4.9.1.2.
When status quo F is applied for 2005 the catch in 2005 is predicted to be 2.38 and 2.30 million tonnes for respectively the deterministic and MCMC projections. These values are quite close to the best guess ( 2 million tonnes) of the WG on the catch level for 2005.

The deterministic projections (based on the SMS maximum likelihood estimate) have a higher initial SSB and lower F compared to the stochastic projections (based on the SMS MCMC estimate). This difference is maintained in the projections such that deterministic projections predict a higher SSB in 2007 for a given TAC in 2006.

Ideally, the two methods should give the same (mean) values for SSB and F. The assessment was made with a lower limit ( 0.25 ) on the CV on catch observations. If this restriction is removed the assessment model give a higher weight on catch observations. This will, as indicated at Figure 4.7.5.7, give a lower SSB and a higher F. The effect of an unbound CV on both types of projections is a relative decrease in predicted SSB. As an example for option a) with a TAC of 2 million tonnes in 2006: SSB in 2007 will in the deterministic projection decrease from 6.0 to 5.7 million tons, whereas the MCMC projection has a decrease from 5.3
to 5.2 million tonnes. Therefore, the difference in the two model outputs is not due to the applied minimum CV.

The confidence limits are likely to be underestimated, because the recruitment estimate for 2005 has been assumed to be without uncertainty. Likewise, the mean weight at age and proportion mature are assumed to be constant.

### 4.9.2 Harvest Control Rules

Presently short term ICES catch advice for blue whiting is based on stochastic predictions which are sensitive to the size of the stock estimated by the most recent assessment. Several assessments methods are carried out on the blue whiting stock. In general they show agreement in the development of the stock and the exploitation but give different estimates of the actual level of stock size and fishing mortality. Often, it is impossible to discriminate that one is better than the other. The difference between the assessments indicates the range where the true situation of the stock and the exploitation of it may be (even though there is no guarantee that the "true" value lies within this range). Short term catch prognoses, therefore, may be very dependent on the choice of a preferred assessment. Also, for a particular assessment method, the results may show large differences from year to year, indicating that it is very sensitive to the weight given to the different data sources used and the addition of new data year.

This may lead to large variation in the short term catch advice which are not only related to variations in the projected fishing mortality and stock size but to a large extent to the variations of the assessment. In such situation the assessment is a poor basis for providing catch options and an alternative procedure should be considered which is less sensitive to the performance of the assessment.

For the credibility of the advice, (large) changes in the catch advice and management measures induced by assessment error are highly undesirable. In such cased it may be preferable to introduce more stability in the annual catch opportunities. HCRs based on a fixed target fishing mortality would not achieve this when this parameter is poorly estimated. Also applying a fixed target F or effort would lead to fluctuations in TACs closely following projections of the projected stock.

In such situations, it could be considered to base management strategy (Harvest Control Rule) on a more robust estimators of the state and the productivity of the stock. This indicator could, for instance, be an index of exploitable biomass obtained from surveys or cpue time series, which are considered to be sufficiently representative indicators for the development of the stock.

As an example, a HCR could be based on a stable (fixed) TAC. In the case of blue whiting, different levels of TAC could be considered associated with observed differences in productivity of the stock. All assessments indicate that between 1995 and 2002 the blue whiting stock has shown a significant increased productivity (high recruitment). This has lead to a large increase of the stock after 1997. Such a situation may justify higher TAC in the period of high stock size caused by increased productivity.

This simple Harvest Control Rule requires only the adoption of a few entities. Firstly a time series should be adopted as relative indicator of exploitable biomass which is the basis for a management decision. It is essential that each year the time series will expanded with a new data point (index). This index is required to decide whether or not to change the management. Secondly, the HCR requires the adoption of a trigger point in the range of the time series. If the index is above the trigger point then an agreed management procedure would be applied. If it is below the trigger point, a different agreed management procedure would be applied. Thirdly it should be agreed what the management procedure would be. In the most simple
situation the management procedure would be a pre-agreed TAC. In principle the rule could be expanded with more trigger points defining ranges within different management procedures would be applied.

Before adopting the HCR, the proposed procedures should be evaluated, with respect to its performance in terms of maintaining the stock within precautionary limits (more here?).

In the case of blue whiting it could be explored whether the acoustic surveys on the spawning grounds could be adopted as representative indicators of the stock size. These surveys have recently been coordinated internationally and expanded by participation of the EU and Far Oer. The survey aims to cover the entire distribution area of blue whiting on the spawning grounds. The surveys are, however, not reliable estimators of expected recruitment. The HCR could be expanded if time series of recruitment became available indicating productivity in the most recent years.

### 4.10 Biological reference points

The present precautionary reference points have been introduced in the advice of ACFM in 1998. The values and their technical basis are:

| Reference <br> point | $\mathbf{B}_{\text {lim }}$ | $\mathbf{B}_{\mathrm{pa}}$ | $\mathbf{F}_{\text {lim }}$ | $\mathbf{F}_{\mathrm{pa}}$ |
| :--- | :--- | :--- | :--- | :--- |
| value | 1.5 mill t | 2.25 mill. t | 0.51 | 0.32 |
| basis | $\mathbf{B}_{\text {loss }}$ | $\mathbf{B}_{\text {lim }} * 1.5$ | $\mathbf{F}_{\text {loss }}$ | $\mathbf{F}_{\text {med }}$ |

Although problems have been identified with these reference points they have remained unchanged since then. A major problem is that fishing at $\mathbf{F}_{\mathrm{pa}}$ implies a high probability of bringing the stock below $\mathbf{B}_{\mathrm{pa}}$, in other words the present combination of $\mathbf{F}_{\mathrm{pa}}$ and $\mathbf{B}_{\mathrm{pa}}$ is inconsistent.

It should be noted that the PA reference points presently applied in the ICES advice are based on an ICA assessment in 1998 and are based on a relatively short time series. Since then regular changes have been made in the assessment by selecting other assessment methods or different assessment configurations. Also major changes have been observed in the stock and in the fishery. The assessments have frequently lead to different perceptions of the history of the development of the stock and the fishing mortality. PA reference points derived from these assessments would vary considerable between years. Since the introduction of the PA reference points the stock has moved into a period of high productivity. In such situations it may be inappropriate to adopt fixed reference points. Therefore, given the uncertaintly in the assessment process, the WG was requested to develop management strategies which are less sensitive to the assessment results.

No attemp was made to recalculate the reference points in 2005. However the value of $\mathbf{B}_{\text {lim }}$ was reconsidered since all management stategies would have to comply with a low probability of bringing the SSB below $\mathbf{B}_{\text {lim }}$.

SGPRP revisited the $\mathbf{B}_{\text {lim }}$ reference point for blue whiting in February 2003 (ICES 2003/ACFM:15). The current $\mathbf{B}_{\text {lim }}$ a value of 1.5 million $t$ was based on an estimate of $\mathbf{B}_{\text {loss }}$ from an assessment in 1998. Since a segmented regression on the stock recruitment data was not significant, $\mathbf{B}_{\text {loss }}$ remains the obvious candidate for $\mathbf{B}_{\text {lim }}$. Within the range of observed SSB there is no indication of reduced recruitment at low SSB. In the assessment carried out in 2002 $\mathbf{B}_{\text {loss }}$ was estimated to be 1.2 million $t$ and SGPRP proposed this value for $\mathbf{B}_{\text {lim }}$.

Based on assessments carried out in 2005 by WGNPBW, the $\mathbf{B}_{\text {loss }}$ is estimated between 1.4 million t (AMCI), 1.7 million t (ISVPA) and 1.6 million t (ICA and SMS), the same values as
last year. If the value of $\mathbf{B}_{\mathrm{lim}}$ is to be revised, it is likely to be around the region of the present value of 1.5 million tonnes.

### 4.11 Management considerations

Countries participating in the fishery on blue whiting have not been able to agree on measures to regulate the fishery. TACs have been set unilateral and increased during the year. Some countries have not set a TAC at all. In the last two years international catches of blue whiting have increased to record high levels. The stock has been able to support these catches because of an increased production since 1995.

No updated, official information on the fishery in 2005 was available. Anectodical information from Working Group members suggest that catches in 2005 may be lower than in 2003 and 2004. It was considered that input into the blue whiting fishery in the second part of the year was somewhat lower than in previous years. The reduced interest in the fishery is likely caused by economic considerations such as an increase in fuel prices and reduced densities of blue whiting in the second part of the year. The Working Group assumed in its forecast that a total catch of 2 million tonnes would be taken in 2005 . This is about 400 thousand tonnes lower than in the previous two years

Should the production change back to the previously observed low-production regime, the current catch levels would be unsustainable., The causes of change in productivity are unknown and need investigation. Maintaining the high catch levels could lead to a rapid decrease or even collapse of the stock in the case of a change to a low productivity regime.

In order to allow optimal harvesting of the stock and to take adequate management measures, monitoring of the productivity of the system should be carried out.

In recent years the fishery has expanded to areas where large amounts of juvenile blue whiting occur. In particular the fishery in the Norwegian Sea has expanded and now contributes about $40 \%$ to the total catches. In order to improve the exploitation, measures should be considered to reduce the exploitation on juveniles.

### 4.12 Quality of the data and the assessment

A new fishery-independent data source, the international ecosystem survey in the Norwegian Sea. was included in this year's assessment. At the same time, the structural uncertainty in the blue whiting assessment has likely increased: while most assessment models assume constant exploitation pattern, the expansion of the fishery in the Norwegian Sea has possibly resulted in increased exploitation of juveniles. This change is not captured in results of any of the assessments. If the change in exploitation pattern has indeed taken place, all assessments are probably too optimistic.

The assessment of blue whiting has been very uncertain in recent years with upwards revision of the stock size with every new assessment. This trend has been driven by exceptionally good recruitment compared to earlier period, while at the same time little information has been available on the recruitment.

Compared to earlier assessments the 2005 assessment shows some upwards revision of the stock from last year; the actual change varies between models (Figure 4.7.6.3), least for ICA and ISVPA but most for AMCI.

The resulting spawning stock in 2004 for the selected models ranges from 5 (AMCI) to 6.6 millions (ISVPA) tonnes and SSB in 2005 from 4.8 to 5.8 million tonnes (Figure 4.7.6.1). As $11 \%$ of age group 1 is mature, estimate of the 2004 yearclass can affect the SSB in 2005. This relatively narrow range of results is however not not indicative of the precision of the
assessement as all the models are somewhat similar (mostly seperable models for fishing mortality). The SMS run where a priori weights on survey relative to catch are varied (figure 4.7.5.4) gives SSB in 2004 in the range from 4 to 7 million tonnes, with a higher number obtained with more weight on surveys. This tendency for the surveys to indicate higher biomass than the catches can also be seen in the ISVPA profiles (4.7.3.1) where the minimum for the catches occur at higher F than the minimum for the Norwegian spawning stock survey that is the most important surveys. Looking at the numbers in figure 4.7.3.1 for the median of the catches and the SSE for the Norwegian spawning stock survey (that are added) shows that the survey will get more weight. Thus, ISVPA is weighting the survey relative to the catch more than the other models with the default settings. An ADAPT model that was run gave spawning stock of 6.8 million tonnes in 2004 but that model has not model for fishing mortality and its fit to the most recent survey is only limited by results from earlier surveys. The perception of the stock today is therefore a question of our confidence in the survey data versus the catch data.

With regard to catch data, fisheries for blue whiting in the eastern Norwegian Sea have expanded. These fisheries are catching also large proportions of young immature fish. This change will violate the assumption of seperability used by most of the assessment models and could affect the assessment results in unpredictable ways.

With regards to the spawning ground survey, figures 4.7.2.2 and 4.7.5.4 show positive residuals for age 5 years and older in the last year, but large negative residuals for ages 2 and 3 years. In Section 4.6.1.2 it was argued that the lack of ages 2 and 3 in the survey 2005 was caused by inadequate sampling as more of those age groups was found in the international survey. The SMS assessment uses age 2 in 2005 from the spawning ground survey for tuning, which is questionable as manifested by the large residual in figure 4.7.5.4; this residual is most likely a sampling problem. Looking at older surveys, negative block is seen in 2001 but the surveyed biomass dropped then by more than $50 \%$ from the years 1999-2000, increasing again after that. Figure 4.12 .1 shows that the survey has generally been a reasonable measure of stock size. Age 5 in 2005 seems to be considerably larger than estimated by the survey and this residual seen there means well over 1.2 million tonnes in spawning stock biomass. Comparing observed and modelled survey biomass seems to indicate that in many of the surveys they fit reasonably together but the age composition in the surveys and the models does not match, indicating sampling problems.

Another measure of the quality of the assessment can be obtained from the estimates of uncertainity from the assesssment models (figures 4.7.2.4, 4.7.2.5, 4.7.3.4, 4.7.5.11, and 4.7.5.13). The uncertainity estimated by ISVPA seems to be considerably more than by AMCI with SMS falling in between. This is caused by ISVPA putting relatively more weight on the noisier survey data but does not have to mean the the ISVPA results are worse. Surveys are often more noisy than data from catches but give the correct main trends ("It is better to be roughly right than to be precisely wrong" - Maynard Keynes). It is known that uncertainty estimated by assessment models is an underestimate of the "real uncertainty" but here the estimated uncertainty seems to explain the difference between assessments.

The selection of a "final run" in an uncertain situation is always a problem but as seen in figures 4.7.5.7 and 4.7.6.1 the selected "final run" is somewhere in the middle of plausible outcomes.

Recruitment estimates have been a problem for this stock with age 1 contributing both to the spawning stock and to the fisheries. Analyses done at this meeting indicated that information from the acoustic surveys in the Norwegian Sea and south of Iceland as well as the bottom trawl surveys in the Barents Sea and Iceland could be used to estimate the size of yearclasses 2003 and 2004, with reasonable confidence (see Section 4.8) .

With regards to yearclasses 2005 and later the question is whether to assume that recruitment has been as in the high production period since 1995 or if to use the whole period since 1981. The geometric mean for the latter period is 35000 million fishes while it is 18000 million fishes for the whole period. Yield per recruit is about 50 g so recruitment as in recent years can sustain a yield close to 1.7 million tonnes while recruitment according to the long term mean can sustain 0.9 million tonnes. Increasing exploitation on age 1 and 2 fish as done in recent years will reduce yield per recruit unless M is considerably higher than $0.2 \mathrm{yr}^{-1}$.

To summarize, the quality of the assessement and recruitment estimates have improved considerably from last year, mostly due to more data on recruitment. As the Norwegian spawning stock and the international ecosystem surveys will likely continue, the assessment might improve further in the coming years. Expansion of the international ecosystem survey in May to include all important nursery areas and better sampling in the spawning stock survey would, however, be beneficial.

### 4.13 Recommendations

- The Working Group recommends merging the data from the international blue whiting spawning stock survey (carried out in 2004 and 2005, and in the future years) with the Norwegian times series in a way that would allow benefiting from the length of the Norwegian time series and from the increased survey effort on the spawning grounds through the international survey. The data should be prepared and evaluated by PGNAPES before the next year.
- The Working Group recognises that more time, tools and expertise is required for developing and evaluating Harvest Control Rules
- The Working Group recognises that several surveys are being carried out in different countries which provide information on blue whiting. Some of the surveys are directed to a mixture of species. The WG request PGNAPES to consider investigating the feasibility of improved coordination between these surveys.
- The Working Group felt that improvement was made in 2005 in getting information on recruitment from blue whiting from existing data sources (surveys). Further work on evaluating existing information sources should be encouraged.
- The Working Group recommends that surveys, which provide information on recruitment of blue whiting should be continued in the same way as in the past. Better coordination of the surveys carried out in different areas is required.
- The Working Group asks ICES to remind member countries that reliable catch statistics are required to enable proper analyses. All countries which have catches of blue whiting should timely report these to ICES. (France did not report catches to ICES in 2004 although catches were reported to the EU. Informal information of Frensh landing came too late to the WG to be included in the assessment.)
- Comparisons of age distributions from landings investigated by different laboratories indicate that there are differences in the age interpretations. Many hardships in the stock assessments may be caused by those uncertainties. The situation is likely ameliorated through the age reading workshop arranged in June 2005. WG recommends that the coordination between the institutes in this line should be continued.
- The Working Group recommends to its members to recompile the time series of catch at age data on an area basis, which allows to analyse the effect of changing
behaviour of the distribution of the fleet. In particular, separate catch at age data for the Norwegian Sea are required.
- The Working Group recommends to the institutes participating in the Norwegian Sea Ecosystem Survey (coordinated by PGNAPES) to make survey data of blue whiting available to PGNAPES in the agreed format to allow for extension of the time series to the years before 2000 .

Table 4.3.1 Landings (tonnes) of BLUE WHITING from the directed fisheries (Sub-areas I and II, Division Va, XIVa and XIVb) 1987-2004, as estimated by the Working Group.

| Country | 1987 | 1988 | $1989{ }^{\text {3) }}$ | 1990 | 1991 | 1992 | 1993 | $1994{ }^{2)}$ | $1995{ }^{3 \text { 3) }}$ | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | - | - | - | - | - | - | - | - | - | - | - | - | 15 | 7,721 | 5,723 | 13,608 | 38,226 | 23,437 |
| Estonia | - | - | - | - | - | - | - | - | - | 377 | 161 | 904 |  |  |  |  |  |  |
| Faroes | 9,290 | - | 1,047 | - | - | - | - | - | - | 345 | - | 44,594 | 11,507 | 17,980 | 64,496 | 82,977 | 115,755 | 109,380 |
| Germany | 1,010 | 3 | 1,341 | - | - | - | - | 2 | 3 | 32 | - | 78 | - | - | 3117 | 1,072 | 813 | 488 |
| Greenland | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |  |  |
| Iceland | - | - | 4,977 | - | - | - | - | - | 369 | 302 | 10,464 | 68,681 ${ }^{\text {4) }}$ | 96,295 | 155,024 | 245,814 | 195,483 | 312,334 | 322,247 |
| Latvia | - | - | - | - | - | - | - | 422 | - | - | - | - | - | - | - | - |  |  |
| Netherlands | - | - | - | - | - | - | - | - | 72 | 25 | - | 63 | 435 | - | 5180 | 906 | 592 | 1,365 |
| Norway ${ }^{5}$ | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 64,581 | 100,922 | 215,075 | 302,166 |
| Norway ${ }^{\text {6 }}$ | - | - | - | 566 | 100 | 912 | 240 | - | - | 58 | 1,386 | 12,132 | 5,455 | - | 28,812 | - |  | 22167 |
| Poland | 56 | 10 | - |  |  | - | - | - | - | - | - | - | - | - | - | - | - |  |

 |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

${ }^{1}$ ) From 1992 only Russia
${ }^{2}$ ) Includes Vb for Russia.
${ }^{3}$ ) Icelandic mixed fishery in Va.
) include mixed in Va and directed in Vb.
${ }^{5}$ ) Directed fishery
${ }^{\circ}$ By-catches of
Table 4.3.2 Landings (tonnes) of BLUE WHITING from directed fisheries (Division Vb,VIa,b, VIIa,b,c and Sub-area XII) 1987-2004, as estimated by the Working Group.

| Country | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | $1998{ }^{\text {1) }}$ | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 2,655 | 797 | 25 | - |  | 3,167 | - | 770 | - | 269 | - | 5051 | 19,625 | 11,856 | 18,110 | 2,141 | 17,813 | 44,992 |
| Estonia |  |  |  | - |  | 6,156 | 1,033 | 4,342 | 7754 | 10,605 | 5,517 | 5,416 |  |  |  |  |  |  |
| Faroes | 70,625 | 79,339 | 70,711 | 43,405 | 10,208 | 12,731 | 14,984 | 22,548 | 26,009 | 18,258 | 22,480 | 26,328 | 93,234 | 129,969 | 188,464 | 115,127 | 208,427 | 206,078 |
| France |  | - | 2,190 |  | - | - | 1,195 | - | 720 | 6,442 | 12,446 | 7,984 | 6,662 | 13,481 | 13,480 | 14,688 | 13,365 | - |
| Germany | 3,850 | 5,263 | 4,073 | 1,699 | 349 | 1,307 | 91 |  | 6,310 | 6,844 | 4,724 | 17,891 | 3,170 | 12,655 | 15,862 | 15,378 | 21,866 | 13,813 |
| Iceland | - | - | - | - | - | - | - | - | - | - | - | - | 64,135 | 105,833 | 119,287 | 91,853 | 189,159 | 99,832 |
| Ireland | 3,706 | 4,646 | 2,014 | - | - | 781 | - | 3 | 222 | 1,709 | 25,785 | 45635 | 35,240 | 25,200 | 29,854 | 17,723 | 22,484 | 62,730 |
| Japan | - | - | - | - | - | 918 | 1,742 | 2,574 | - | - | - | - |  | - | - | - |  |  |
| Latvia | - | - | - | - | - | 10,742 | 10,626 | 2,160 | - | - | - | - | - | - | - | - | - |  |
| Lithauen | - | - | - | - | - |  | 2,046 | - | - | - | - | - | - | - | - | - | - |  |
| Netherlands ${ }^{2}$ ) | 5,627 | 800 | 2,078 | 7,280 | 17,359 | 11,034 | 18,436 | 21,076 | 26,703 | 17,644 | 23,676 | 27,884 | 35,408 | 46,128 | 68,415 | 33,365 | 45,239 | 82,520 |
| Norway | 191,012 | 208,416 | 258,386 | 281,036 | 114,866 | 148,733 | 198,916 | 226,235 | 261,272 | 337,434 | 318,531 | 519,622 | 475,004 | 460,274 | 399,932 | 385,495 | 502,320 | 486,843 |
| UK (Scotland) | 3,315 | 5,071 | 8,020 | 6,006 | 3,541 | 6,849 | 2,032 | 4,465 | 10,583 | 14,325 | 33,398 | 92,383 | 98,853 | 42,478 | 50,147 | 26,403 | 27,136 | 56,326 |
| Sweden |  |  |  |  |  |  |  |  | - | - |  | - |  | - |  | 10 | - |  |
| USSR/Russia ${ }^{3}$ ) | 165,497 | 121,705 | 127,682 | 124,069 | 72,623 | 115,600 | 96,000 | 94,531 | 83,931 | 64,547 | 68,097 | 79,000 | 112,247 | 141,257 | 141,549 | 144,419 | 163,812 | 179,400 |
| Total | 446,287 | 426,037 | 475,179 | 463,495 | 218,946 | 318,018 | 347,101 | 378,704 | 423,504 | 478,077 | 514,654 | 827,194 | 943,578 | 989,131 | 1,045,100 | 846,602 | 1,211,621 | 1,232,534 |
| ${ }^{1}$ ) Including some directed fishery also in Division IVa. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{2}$ ) Revised for the years 1987, 1988, 1989, 1992, 1995, 1996, 1997 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{3}$ ) From 1992 only Russia |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{4)}$ Reported to the EU but not to the ICES WGNPBW. (Landings of 19,467 tonnes) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 4.3.3 Landings (tonnes) of BLUE WHITING from directed fisheries and by-catches caught in other fisheries (Divisions IIIa, IV) 1987-2004, as estimated by he Working Group.

| Country | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | $1993{ }^{3)}$ | 1994 | 1995 | 1996 | 1997 | $1998{ }^{2)}$ | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark ${ }^{4)}$ |  |  | 3,632 | 10,972 | 5,961 | 4,438 | 25,003 | 5,108 | 4,848 | 29,137 | 9,552 | 40,143 | 36,492 | 30,360 | 21,995 |  |  |  |
| Denmark ${ }^{5}$ | 28,541 | 18,144 | 22,973 | 16,080 | 9,577 | 26,751 | 16,050 | 14,578 | 7,591 | 22,695 | 16,718 | 16,329 | 8,521 | 7,749 | 7,505 | 35,530 | 26,896 | 21,071 |
| Faroes ${ }^{4 / 6)}$ |  |  |  |  |  |  |  |  |  |  |  | - | - | - | 60 |  |  |  |
| Faroes ${ }^{5 / 6)}$ | 7,051 | 492 | 3,325 | 5,281 | 355 | 705 | 1,522 | 1,794 | - | 6,068 | 6,066 | 296 | 265 | 42 | 6,741 | 7,317 | 5,712 | 6,864 |
| Germany ${ }^{1}$ | 115 | 280 | 3 | - | - | 25 | 9 | - | - | - | - |  |  | - | 81 | - | 36 | 19 |
| Ireland | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 4 |  | 4 |
| Netherlands | - | - | - | 20 | - | 2 | 46 | - | - | - | 793 |  |  | - | - | 50 | 0 | 0 |
| Norway ${ }^{4}$ | 24,969 | 24,898 | 42,956 | 29,336 | 22,644 | 31,977 | 12,333 | 3,408 | 78,565 | 57,458 | 27,394 | 28,814 | 48,338 | 73,006 | 21,804 | 85,062 | 117,145 | 107,311 |
| Norway ${ }^{5}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 58,182 |  |  |  |
| Russia | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 69 | - | - |  |
| Scotland |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 35 |
| Sweden | 2,013 | 1,229 | 3,062 | 1,503 | 1,000 | 2,058 | 2,867 | 3,675 | 13,000 | 4,000 | 4,568 | 9,299 | 12,993 | 3,319 | 2,086 | 17,689 | 8,326 | 3,289 |
| UK | - | 100 | 7 | - | 335 | 18 | 252 | - | - | 1 | - | - | - | - | - | - | 65 |  |
| Total | 62,689 | 45,143 | 75,958 | 63,192 | 39,872 | 65,974 | 58,082 | 28,563 | 104,004 | 119,359 | 65,091 | 94,881 | 106,609 | 114,476 | 118,523 | 145,652 | 158,180 | 138,593 |
| ${ }^{1}$ ) Including directed fishery also in Division IVa. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{2}$ ) Including mixed industrial fishery in the Norwegian Sea |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{3}$ ) Imprecise estimates for Sweden: reported catch of 34265 t in 1993 is replaced by the mean of 1992 and 1994 , i.e. $2,867 \mathrm{t}$, and used in the assessment. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{4}$ Directed fishery |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{5}$ ) By -catches of blue whiting in other fisheries. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{6}$ For the periode 1987-2000 landings figures also include landings from mixed fisheries in Division Vb . |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 4.3.4 Landings (tonnes) of BLUE WHITING from the Southern areas (Sub-areas VIII and IX and Divisions VIIg-k and VIId,e) 1987-2004, as estimated by the Working Group.

| Country | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Germany | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | $600{ }^{2)}$ | $88^{2)}$ | 973 |
| Ireland | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | $98^{2)}$ | $96^{2)}$ | 12,659 |
| Netherlands | - | - | - | 450 | 10 | - | - | - | - | - | - | $10^{1)}$ | - |  | - | $3208{ }^{2)}$ | 2471, ${ }^{2}$ | 11,426 |
| Norway | 4 | - | - | - | - | - | - | - | - | - | - |  |  | - | - | - | - | 39197 |
| Portugal | 9,148 | 5,979 | 3,557 | 2,864 | 2,813 | 4,928 | 1,236 | 1,350 | 2,285 | 3,561 | 2,439 | 1,900 | 2,625 | 2,032 | 1,746 | 1,659 | 2,651 | 3,937 |
| Russia |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 685 |
| Scotland |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 603 |
| Spain | 23,644 | 24,847 | 30,108 | 29,490 | 29,180 | 23,794 | 31,020 | 28,118 | 25,379 | 21,538 | 27,683 | 27,490 | 23,777 | 22,622 | 23,218 | 17,506 | 13,825 | 15,612 |
| UK | 23 | 12 | 29 | 13 | - | - | - | 5 | - | - | - | - | - | - | - | - | 181 |  |
| France | - | - | 1 | - | - | - | - | - | - | - | - | - | - | - | - | - | 784 |  |
| Total | 32,819 | 30,838 | 33,695 | 32,817 | 32,003 | 28,722 | 32,256 | 29,473 | 27,664 | 25,099 | 30,122 | 29,400 | 26,402 | 24,654 | 24,964 | 23,071 | 20,097 | 85,093 |

Table 4.3.5 Total landings of blue whiting by country and area for 2004 in tonnes. Landing figures provided by Working Group members and these figures may not be official catch statistics and therefore can not be used for management purposes.

|  |  |  |  |  |  |  |  |  |  |  |  | $8$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| I |  |  |  |  |  |  | 63 |  |  |  |  |  |  | 63 |
| IIa | 23,437 | 95,868 |  | 386 | 183,322 |  | 314,690 |  | 137,430 | 64 |  | 15,794 | 1,365 | 772,355 |
| IIb |  |  |  | 103 | 392 |  | 591 |  | 28,976 |  |  |  |  | 30,062 |
| IIIa | 4,274 | 53 |  |  |  |  | 383 |  |  |  |  | 2,730 |  | 7,440 |
| IVa | 16,368 | 6,627 |  | 19 |  | 4 | 106,344 |  |  | 35 |  | 532 | 0 | 129,929 |
| IVb | 429 | 184 |  |  |  |  | 584 |  |  |  |  | 27 | 0 | 1,224 |
| IXa |  |  |  |  |  |  | 0 | 3,937 |  |  |  |  |  | 3,937 |
| Va |  | 13,512 |  |  | 138,533 |  | 8,989 |  |  |  |  |  |  | 161,034 |
| Vb | 12,935 | 111,036 |  | 395 | 95,090 | 1,653 | 18,790 |  | 104,371 | 1,364 |  |  | 3,143 | 348,777 |
| VIa | 31,935 | 44,632 |  | 13,196 |  | 42,506 | 67,890 |  |  | 53,587 |  |  | 62,944 | 316,690 |
| VIb |  |  |  |  | 4,742 |  | 320,364 |  | 69,096 |  |  |  |  | 394,202 |
| VIIa |  |  |  |  |  | 0 | 0 |  |  |  |  |  |  | 0 |
| VIIb |  |  |  | 2 |  | 1,524 | 0 |  |  | 1,376 |  |  | 140 | 3,042 |
| VIIc | 122 | 47,247 |  | 220 |  | 15,538 | 69,616 |  | 871 |  |  |  | 16,293 | 149,906 |
| VIIg |  |  |  |  |  | 4 | 0 |  |  |  |  |  |  | 4 |
| VIIIabd |  |  |  |  |  |  | 0 |  |  |  |  |  | 131 | 131 |
| VIIIC+IXa |  |  |  |  |  |  | 0 |  |  |  | 15,612 |  |  | 15,612 |
| VIIj |  |  |  | 31 |  | 925 | 0 |  |  | 603 |  |  | 895 | 2,454 |
| VIIk |  |  |  | 942 |  | 11,730 | 39,197 |  | 685 |  |  |  | 10,400 | 62,955 |
| XII |  | 3,163 |  |  |  | 1,509 | 10,183 |  | 5,062 |  |  |  |  | 19,917 |
| XIVb |  |  |  |  |  |  | 0 |  | 271 |  |  |  |  | 271 |
| Grand Total | 89,500 | 322,322 | $1)$ | 15,293 | 422,079 | 75,393 | 957,684 | 3,937 | 346,762 | 57,028 | 15,612 | 19,083 | 95,311 | 2,420,005 |

[^1]Table 4.3.6 Landings (tonnes) of BLUE WHITING from the main fisheries, 1987-2004, as estimated by the Working Group.

| Area | Norwegian Sea fishery <br> (Sub-areas $1+2$ and <br> Divisions Va, XIVa-b) | Fishery in the spawning area (Divisions Vb, VIa, VIb and VIIb-c) | Directed- and mixed fisheries (Divisions IIII and IV ) | Total northern areas | Total southern areas (Subareas VIII and IX and Divisions VIId, e, $\mathrm{g}-\mathrm{k})$ | Grand total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 123,042 | 446,287 | 62,689 | 632,018 | 32,819 | 664,837 |
| 1988 | 55,829 | 426,037 | 45,143 | 527,009 | 30,838 | 557,847 |
| 1989 | 42,615 | 475,179 | 75,958 | 593,752 | 33,695 | 627,447 |
| 1990 | 2,106 | 463,495 | 63,192 | 528,793 | 32,817 | 561,610 |
| 1991 | 78,703 | 218,946 | 39,872 | 337,521 | 32,003 | 369,524 |
| 1992 | 62,312 | 318,081 | 65,974 | 446,367 | 28,722 | 475,089 |
| 1993 | 43,240 | 347,101 | 58,082 | 448,423 | 32,256 | 480,679 |
| 1994 | 22,674 | 378,704 | 28,563 | 429,941 | 29,473 | 459,414 |
| 1995 | 23,733 | 423,504 | 104,004 | 551,241 | 27,664 | 578,905 |
| 1996 | 23,447 | 478,077 | 119,359 | 620,883 | 25,099 | 645,982 |
| 1997 | 62,570 | 514,654 | 65,091 | 642,315 | 30,122 | 672,437 |
| 1998 | 177,494 | 827,194 | 94,881 | 1,099,569 | 29,400 | 1,128,969 |
| 1999 | 179,639 | 943,578 | 106,609 | 1,229,826 | 26,402 | 1,256,228 |
| 2000 | 284,666 | 989,131 | 114,477 | 1,388,274 | 24,654 | 1,412,928 |
| 2001 | 591,583 | 1,045,100 | 118,523 | 1,755,206 | 24,964 | 1,780,170 |
| 2002 | 541,467 | 846,602 | 145,652 | 1,533,721 | 23,071 | 1,556,792 |
| 2003 | 931,508 | 1,211,621 | 158,180 | 2,301,309 | 20,097 | 2,321,406 |
| 2004 | 963,785 | 1,232,534 | 138,593 | 2,334,912 | 85,093 | 2,420,005 ${ }^{10}$ |

${ }^{1)}$ Preliminary data, not including France. Iceland revised their data after the meeting of 42 436. In the assessment was used 2,377,569.

Table 4.3.7 Total landings of blue whiting by quarter and area for 2004 in tonnes. Landing figures provided by Working Group members.

| Area | Quarter 1 | Quarter 2 | Quarter 3 | Quarter 4 | Grand Total |  |
| :--- | ---: | ---: | ---: | ---: | ---: | :---: |
| I |  |  |  |  |  |  |
| IIIa | 13,776 | 294,494 | 355,895 | 108,190 | 772,355 |  |
| IIb |  | 455 | 1,512 | 28,095 | 30,062 |  |
| IIIa | 13 | 528 | 6,816 | 83 | 7,440 |  |
| IVa | 26,917 | 37,229 | 40,785 | 24,998 | 129,929 |  |
| IVb | 3 | 20 | 1,138 | 63 | 1,224 |  |
| IXa | 606 | 920 | 768 | 1,643 | 3,937 |  |
| Va | 682 | 56,669 | 55,745 | 47,938 | 161,034 |  |
| Vb | 20,863 | 250,079 | 3,401 | 74,434 | 348,777 |  |
| VIa | 136,186 | 178,310 | 1,621 | 574 | 316,690 |  |
| VIb | 338,240 | 55,215 | 747 |  | 394,202 |  |
| VIIa |  |  |  | 0 | 0 |  |
| VIIb | 2,475 | 567 | 0 |  | 3,042 |  |
| VIIc | 149,342 | 564 |  |  | 149,906 |  |
| VIIg |  | 4 | 0 | 0 | 4 |  |
| VIIIabd |  | - | 131 |  | 131 |  |
| VIIIc+IXa | 4,295 | 3,600 | 3,599 | 4,117 | 15,612 |  |
| VIIj | 1,209 | 1,166 | 6 | 72 | 2,454 |  |
| VIIk | 62,955 |  |  |  | 62,955 |  |
| XII | 19,917 |  | 271 |  | 19,917 |  |
| XIVb | 777,480 | 880,092 | 472,227 | 290,206 | $2,420,005$ |  |
| Grand Total |  |  |  |  | 271 |  |

Table 4.5.1.1. Blue whiting. Total landings, No. of samples, No. of fish measured and No. of fish aged by country and quarter for 2004.

| Country | Quarter | Landings (t) | No. Samples | No. Fish measured | No. Fish aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 1 | 21842 | 10 | 805 | 778 |
|  | 2 | 40973 | 8 | 495 | 454 |
|  | 3 | 22236 | 14 | 819 | 817 |
|  | 4 | 4449 | 1 | 100 | 50 |
|  | Total | 89500 | 33 | 2219 | 2099 |
| Faroe Islands | 1 | 96047 | 12 | 2417 | 1200 |
|  | 2 | 146734 | 10 | 2309 | 1000 |
|  | 3 | 37915 | 5 | 1146 | 500 |
|  | 4 | 41626 | 4 | 750 | 400 |
|  | Total | 322322 | 31 | 6622 | 3100 |
| Germany | 1 | 5696 | 0 | 0 | 0 |
|  | 2 | 9538 | 0 | 0 | 0 |
|  | 3 | 59 | 0 | 0 | 0 |
|  | 4 | 0 | 0 | 0 | 0 |
|  | Total | 15293 | 0 | 0 | 0 |
| Iceland | 1 | 4599 | 1 | 100 | 50 |
|  | 2 | 223594 | 47 | 3629 | 2310 |
|  | 3 | 94881 | 17 | 1143 | 843 |
|  | 4 | 56569 | 11 | 690 | 518 |
|  | Total | 379643 | 76 | 5562 | 3721 |
| Ireland | 1 | 52885 | 13 | 2246 | 1248 |
|  | 2 | 21914 | 0 | 0 | 0 |
|  | 3 | 517.576 | 0 | 0 | 0 |
|  | 4 | 76 | 0 | 0 | 0 |
|  | Total | 75393 | 13 | 2246 | 1248 |
| Norway | 1 | 442937 | 176 | 6783 | 3855 |
|  | 2 | 256421 | 205 | 8745 | 5925 |
|  | 3 | 232108 | 125 | 6345 | 1483 |
|  | 4 | 26218 | 51 | 1953 | 151 |
|  | Total | 957684 | 557 | 23826 | 11414 |
| Portugal | 1 | 606 | 79 | 8960 | 321 |
|  | 2 | 920 | 85 | 10781 | 262 |
|  | 3 | 768 | 74 | 8105 | 162 |
|  | 4 | 1643 | 75 | 9268 | 162 |
|  | Total | 3937 | 313 | 37114 | 907 |
| Russia | 1 | 66065 | 108 | 27407 | 1066 |
|  | 2 | 107466 | 27 | 5031 | 555 |
|  | 3 | 61172 | 105 | 17512 | 813 |
|  | 4 | 112059 | 131 | 23294 | 655 |
|  | Total | 346762 | 371 | 73244 | 3089 |
| Scotland | 1 | 31252 | 1 | 151 | 135 |
|  | 2 | 25472 | 3 | 662 | 186 |
|  | 3 |  | 0 | 0 | 0 |
|  | 4 | 305 | 0 | 0 | 0 |
|  | Total | 57028 | 4 | 813 | 321 |
| Spain | 1 | 4295 | 69 | 6850 | 40 |
|  | 2 | 3600 | 74 | 6559 | 45 |
|  | 3 | 3599 | 78 | 6605 | 22 |
|  | 4 | 4117 | 83 | 7776 | 30 |
|  | Total | 15612 | 304 | 27790 | 137 |
| Sweden | 1 | 0 | 0 | 0 | 0 |
|  | 2 | 379 | 0 | 0 | 0 |
|  | 3 | 18682 | 0 | 0 | 0 |
|  | 4 | 22 | 0 | 0 | 0 |
|  | Total | 19083 | 0 | 0 | 0 |
| The Netherlands | 1 | 51254 | 68 | 1699 | 1699 |
|  | 2 | 43081 | 4 | 100 | 100 |
|  | 3 | 289 | 0 | 0 | 0 |
|  | 4 | 687 | 0 | 0 | 0 |
|  | Total | 95311 | 72 | 1799 | 1799 |
| Grand Total |  | 2377569 | 1774 | 181235 | 27835 |

Table 4.5.1.2 Blue Whiting. Sampling levels in 2004 per area.

| Area | landings | nos samples | nos measured | nos aged |
| :---: | :---: | :---: | :---: | :---: |
| I | 63 | 0 | 0 | 0 |
| IIa | 772,355 | 410 | 41,835 | 8,239 |
| IIb | 30,062 | 95 | 15,471 | 642 |
| IIIa | 7,440 | 9 | 548 | 496 |
| IVa | 129,929 | 186 | 8,691 | 2,014 |
| IVb | 1,224 | 2 | 149 | 49 |
| IXa | 3,937 | 313 | 37,114 | 907 |
| Va | 118,598 | 33 | 3,208 | 1,815 |
| Vb | 348,777 | 55 | 6,019 | 2,052 |
| VIa | 316,690 | 101 | 5,101 | 4,589 |
| VIb | 394,202 | 168 | 27,449 | 2,942 |
| VIIa |  | 0 | 0 | 0 |
| VIIb | 3,042 | 3 | 305 | 255 |
| VIIc | 149,906 | 48 | 2,870 | 2,413 |
| VIIg |  | 0 | 0 | 0 |
| VIIIabd | 131 | 0 | 0 | 0 |
| VIIIc+IXa | 15,612 | 304 | 27,790 | 137 |
| VIIj | 2,454 | 0 | 0 | 0 |
| VIIk | 62,955 | 35 | 2,413 | 710 |
| XII | 19,917 | 12 | 2,272 | 575 |
| XIVb | 271 | 0 | 0 | 0 |
| total | 2,377,569 | 1,774 | 181,235 | 27,835 |

Table 4.5.4.1 Blue whiting. Landing in numbers ('000) by length group (cm) and quarters for the Nothern area in 2004.

| Length | Q1 | Q2 | Q3 | Q4 | All year |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 5 |  |  |  |  |  |
| 6 |  |  |  |  |  |
| 7 |  |  |  |  |  |
| 8 |  |  |  |  |  |
| 9 |  | 2256 |  |  | 2256 |
| 10 |  |  | 146 |  | 146 |
| 11 |  |  | 3342 |  | 3342 |
| 12 |  |  | 2664 |  | 2664 |
| 13 | 13 | 783 | 1747 |  | 2543 |
| 14 | 466 | 391 | 1529 | 34 | 2420 |
| 15 | 2948 | 329 | 1857 | 578 | 5712 |
| 16 | 8671 | 2349 | 1201 | 306 | 12527 |
| 17 | 15394 | 10906 | 6318 | 952 | 33570 |
| 18 | 12900 | 35810 | 30874 | 12541 | 92125 |
| 19 | 9057 | 101520 | 139518 | 36966 | 287060 |
| 20 | 27842 | 111267 | 164576 | 114534 | 418218 |
| 21 | 36487 | 127270 | 189820 | 110739 | 464316 |
| 22 | 70790 | 304988 | 360062 | 100157 | 835998 |
| 23 | 213974 | 616330 | 518325 | 116331 | 1464960 |
| 24 | 502694 | 918476 | 593181 | 139884 | 2154235 |
| 25 | 767692 | 1040007 | 594793 | 224068 | 2626560 |
| 26 | 784481 | 949343 | 478809 | 233577 | 2446210 |
| 27 | 723627 | 822375 | 367655 | 247984 | 2161642 |
| 28 | 595500 | 580860 | 239297 | 168652 | 1584308 |
| 29 | 450614 | 433268 | 147004 | 86406 | 1117291 |
| 30 | 335525 | 317997 | 107085 | 88275 | 848882 |
| 31 | 230597 | 171218 | 68989 | 53090 | 523895 |
| 32 | 148296 | 126876 | 24880 | 40579 | 340632 |
| 33 | 114727 | 81950 | 9797 | 18490 | 224964 |
| 34 | 56576 | 21116 | 6082 | 6651 | 90425 |
| 35 | 33223 | 20483 | 1663 | 2552 | 57920 |
| 36 | 17603 | 23688 | 918 | 1224 | 43432 |
| 37 | 11554 | 8791 | 1410 | 1960 | 23715 |
| 38 | 4873 | 3084 |  | 986 | 8942 |
| 39 | 3961 |  |  | 306 | 4267 |
| 40 | 2779 | 263 | 36 | 102 | 3181 |
| 41 | 1633 |  |  | 68 | 1701 |
| 42 | 1080 |  |  | 34 | 1114 |
| 43 | 377 |  |  | 34 | 411 |
| 44 | 38 |  |  |  | 38 |
| 45 | 38 |  |  | 34 | 72 |
| 46 |  |  |  |  |  |
| 47 | 170 |  |  |  | 170 |
| 48 |  |  |  |  |  |
| 49 | 170 |  |  |  | 170 |
| 50 |  |  |  |  |  |
| TOTAL numbers | 5186371 | 6833993 | 4063578 | 1808090 | 17892031 |

Table 4.5.4.2 Blue whiting. Landings in numbers ('000) by length group (cm) and quarters for the North Sea and Skagerrak in 2004.

| Length | Q1 | Q2 | Q3 | Q4 | All year |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 5 |  |  |  |  |  |
| 6 |  |  |  |  |  |
| 7 |  |  |  |  |  |
| 8 |  |  | 356 |  | 356 |
| 9 | 281 |  |  |  | 281 |
| 10 | 281 |  |  |  | 281 |
| 11 | 281 |  |  | 455 | 736 |
| 12 |  |  |  | 303 | 303 |
| 13 | 281 | 380 |  | 303 | 964 |
| 14 | 5414 | 1902 |  | 4704 | 12020 |
| 15 | 34466 | 6032 | 6400 | 6373 | 53271 |
| 16 | 66878 | 25165 | 9955 | 6980 | 108978 |
| 17 | 81653 | 44530 | 2848 | 6980 | 136011 |
| 18 | 45747 | 61196 | 17150 | 4552 | 128645 |
| 19 | 22453 | 84637 | 46399 | 7586 | 161075 |
| 20 | 18985 | 73711 | 75846 | 32318 | 200860 |
| 21 | 35082 | 63459 | 73151 | 42332 | 214024 |
| 22 | 33398 | 85882 | 101506 | 32925 | 253711 |
| 23 | 26382 | 76172 | 108674 | 29435 | 240663 |
| 24 | 13091 | 61919 | 86063 | 24580 | 185653 |
| 25 | 10384 | 46219 | 67787 | 23821 | 148211 |
| 26 | 10946 | 28751 | 40205 | 19725 | 99627 |
| 27 | 5794 | 18343 | 22143 | 14111 | 60391 |
| 28 | 8221 | 8696 | 16664 | 9407 | 42988 |
| 29 | 5595 | 6139 | 8848 | 7435 | 28017 |
| 30 | 2707 | 3381 | 4858 | 5766 | 16712 |
| 31 | 1403 | 2620 | 3286 | 3186 | 10495 |
| 32 |  | 792 | 1410 | 2731 | 4933 |
| 33 | 1123 | 517 | 783 | 1062 | 3485 |
| 34 | 561 | 137 | 39 | 759 | 1496 |
| 35 |  | 137 | 14 | 607 | 758 |
| 36 |  |  | 721 | 152 | 873 |
| 37 | 281 | 275 | 719 |  | 1275 |
| 38 |  |  | 362 |  | 362 |
| 39 |  |  |  |  |  |
| 40 |  |  | 2 |  | 2 |
| 41 |  |  | 711 |  | 711 |
| 42 |  |  |  |  |  |
| 43 |  |  |  | 152 | 152 |
| 44 |  |  |  |  |  |
| 45 |  |  |  |  |  |
| 46 |  |  |  |  |  |
| 47 |  |  |  |  |  |
| 48 |  |  |  |  |  |
| 49 |  |  |  |  |  |
| 50 |  |  |  |  |  |
| TOTAL numbers | 431688 | 700992 | 696900 | 288740 | 2118320 |

Table 4.5.4.3 Blue whiting. Landings in numbers ('000) by length group (cm)
and quarters for the Southern area in 2004.

| Length | Q1 | Q2 | Q3 | Q4 | All year |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 5 |  |  |  |  |  |
| 6 |  |  |  |  |  |
| 7 |  |  |  |  |  |
| 8 |  |  |  |  |  |
| 9 |  |  |  |  |  |
| 10 |  |  |  |  |  |
| 11 |  |  |  |  |  |
| 12 |  |  |  |  |  |
| 13 |  |  |  |  |  |
| 14 |  |  |  |  |  |
| 15 | 122 | 1 | 20 |  | 143 |
| 16 | 1399 | 22 | 72 | 28 | 1521 |
| 17 | 2569 | 504 | 391 | 81 | 3546 |
| 18 | 3744 | 2804 | 2433 | 1424 | 10406 |
| 19 | 7344 | 6785 | 4891 | 6847 | 25866 |
| 20 | 8520 | 12460 | 8547 | 12215 | 41742 |
| 21 | 8442 | 12713 | 15368 | 13461 | 49983 |
| 22 | 8373 | 11324 | 14244 | 12156 | 46097 |
| 23 | 7128 | 6664 | 8559 | 9851 | 32202 |
| 24 | 6428 | 5681 | 4551 | 6561 | 23222 |
| 25 | 4129 | 3247 | 1774 | 3614 | 12764 |
| 26 | 2755 | 1855 | 1757 | 2339 | 8706 |
| 27 | 1746 | 784 | 635 | 1274 | 4440 |
| 28 | 1715 | 722 | 436 | 686 | 3558 |
| 29 | 1059 | 275 | 163 | 324 | 1821 |
| 30 | 792 | 232 | 106 | 248 | 1377 |
| 31 | 805 | 100 | 58 | 147 | 1110 |
| 32 | 333 | 85 | 38 | 69 | 525 |
| 33 | 215 | 20 | 20 | 56 | 311 |
| 34 | 61 | 8 | 7 | 19 | 95 |
| 35 | 36 | 9 | 4 | 11 | 60 |
| 36 | 32 | 4 | 1 | 7 | 43 |
| 37 | 7 | 2 |  | 5 | 14 |
| 38 |  |  |  | 7 | 7 |
| 39 |  |  |  | 2 | 3 |
| 40 |  | 1 |  | 2 | 3 |
| 41 |  |  |  |  |  |
| 42 |  |  |  |  |  |
| 43 |  |  |  |  |  |
| 44 |  |  |  |  |  |
| 45 |  |  |  |  |  |
| 46 |  |  |  |  |  |
| 47 |  |  |  |  |  |
| 48 |  |  |  |  |  |
| 49 |  |  |  |  |  |
| 50 |  |  |  |  |  |
| TOTAL numbers | 67756 | 66300 | 64076 | 71434 | 269566 |

Table 4.5.4.4 BLUE WHITING. Catch in number (millions) by age group in the directed fisheries (Sub-areas I and II, Divisions Va, and XIVa+b, Vb, VIa+b, VIIbc and VIIg-k) in 1993-2004.

| Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | - | - | 1 | 4 | 167 | 15 | 61 | 41 | 119 | 16 | 58 | 6 |
| 1 | 37 | 44 | 99 | 497 | 1352 | 984 | 544 | 912 | 3459 | 1111 | 2464 | 1132 |
| 2 | 130 | 31 | 143 | 327 | 1079 | 3535 | 1180 | 752 | 3924 | 2439 | 3626 | 3481 |
| 3 | 335 | 190 | 338 | 451 | 751 | 3211 | 5257 | 3119 | 2728 | 2939 | 7964 | 6220 |
| 4 | 1348 | 362 | 416 | 425 | 526 | 929 | 3235 | 4834 | 3644 | 2114 | 4726 | 6524 |
| 5 | 376 | 1242 | 566 | 248 | 268 | 346 | 362 | 1517 | 2474 | 1804 | 2006 | 2972 |
| 6 | 196 | 294 | 769 | 430 | 238 | 311 | 186 | 500 | 555 | 1602 | 1090 | 1252 |
| 7 | 108 | 201 | 246 | 619 | 270 | 298 | 143 | 210 | 160 | 336 | 398 | 633 |
| 8 | 60 | 103 | 154 | 214 | 391 | 257 | 146 | 144 | 91 | 165 | 119 | 246 |
| 9 | 38 | 88 | 58 | 88 | 101 | 209 | 66 | 57 | 69 | 100 | 18 | 74 |
| $10+$ | 14 | 32 | 40 | 70 | 164 | 85 | 138 | 139 | 55 | 142 | 27 | 36 |
| Total | 2,641 | 2,588 | 2,829 | 3,373 | 5,307 | 10,180 | 11,318 | 12,225 | 17,281 | 12,768 | 22,495 | 22,575 |

Tonnes $389,010401,378447,015493,373545,0581,000,8701,123,3171,273,1231,636,6831,399,6592,177,0472,219,296$

Table 4.5.4.5 BLUE WHITING. Catch in number (million) by age group in the directed fishery and bycatches from mixed fisheries (Divisions IIIa and IV) for 1993-2004.

| Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 132 | 95 | 3303 | 812 | 29 | 11 | 60 | 56 | 9 | 190 | 222 | 52 |
| 1 | 167 | 33 | 101 | 1334 | 621 | 576 | 188 | 822 | 770 | 621 | 1191 | 925 |
| 2 | 39 | 21 | 88 | 71 | 269 | 524 | 286 | 317 | 416 | 685 | 369 | 784 |
| 3 | 91 | 18 | 29 | 58 | 50 | 259 | 434 | 253 | 174 | 274 | 368 | 405 |
| 4 | 97 | 37 | 11 | 71 | 14 | 47 | 168 | 143 | 149 | 105 | 73 | 116 |
| 5 | 15 | 6 | 6 | 39 | 14 | 6 | 16 | 22 | 109 | 17 | 18 | 46 |
| 6 | 7 | 3 | 11 | 45 | 5 | 4 | 5 | 3 | 29 | 45 | 23 | 12 |
| 7 | 8 | 1 | 2 | 33 | 4 | 3 | 5 | 0 | 9 | 8 | 1 | 11 |
| 8 | - | 1 | 2 | 14 | 6 | 4 | 6 | 7 | 6 | 3 | 1 | 1 |
| 9 | - | 0 | 1 | 9 | 1 | 4 | 1 | 1 | 8 | 2 | 1 | 1 |
| $10+$ | - | - | 1 | 11 | 2 | 12 | 3 | 1 | 11 | 1 | 1 | 1 |
| Total | 556 | 214 | 3,555 | 2,499 | 1,015 | 1,450 | 1,172 | 1,627 | 1,689 | 1,951 | 2,269 | 2,355 |
| Tonnes 55,215 | 28,563 | $104,004119,359$ | 65,091 | 94,881 | 106,609 | 114,477 | 118,523 | 136,171 | 153,697 | 138,593 |  |  |

Table 4.5.4.6 BLUE WHITING. Catch in number (millions) by age group in the Southern area, 1993-2004.

| Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 25 | 13 | 3 | 9 | 11 | 18 | 18 | 32 | 33 | 17 | 7 | 4 |
| 1 | 41 | 12 | 96 | 43 | 118 | 97 | 57 | 80 | 134 | 88 | 88 | 84 |
| 2 | 146 | 56 | 123 | 131 | 143 | 122 | 82 | 123 | 146 | 108 | 79 | 130 |
| 3 | 181 | 149 | 55 | 117 | 86 | 71 | 130 | 93 | 60 | 79 | 47 | 50 |
| 4 | 62 | 72 | 38 | 36 | 26 | 69 | 57 | 35 | 14 | 24 | 26 | 10 |
| 5 | 12 | 27 | 44 | 33 | 8 | 32 | 35 | 9 | 10 | 4 | 12 | 5 |
| 6 | 7 | 9 | 20 | 17 | 4 | 7 | 15 | 10 | 1 | 1 | 4 | 3 |
| 7 | 2 | 5 | 6 | 5 | 3 | 2 | 3 | 3 | 0 | 0 | 1 | 1 |
| $8+$ | 1 | 4 | 5 | 3 | 3 | 4 | 2 | 0 | 0 | 0 | 1 | 0 |
| Total | 477 | 347 | 390 | 394 | 402 | 422 | 399 | 384 | 398 | 321 | 264 | 286 |
| Tonnes 32,256 | 29,468 | 27,664 | 25,099 | 30,122 | 29,400 | 26,402 | 24,654 | 24,964 | 19,165 | 16,476 | 19,680 |  |

Table 4.5.4.7. Blue Whiting: Catch in numbers (thousands) of the total stock in 1981-2004.

| Age |  | $\mathbf{1 9 8 1}$ | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{1}$ | 258,000 | 148,000 | $2,283,000$ | $2,291,000$ | $1,305,000$ | 650,000 | 838,000 |
|  | $\mathbf{2}$ | 348,000 | 274,000 | 567,000 | $2,331,000$ | $2,044,000$ | 816,000 | 578,000 |
|  | $\mathbf{3}$ | 681,000 | 326,000 | 270,000 | 455,000 | $1,933,000$ | $1,862,000$ | 728,000 |
|  | $\mathbf{4}$ | 334,000 | 548,000 | 286,000 | 260,000 | 303,000 | $1,717,000$ | $1,897,000$ |
|  | $\mathbf{5}$ | 548,000 | 264,000 | 299,000 | 285,000 | 188,000 | 393,000 | 726,000 |
|  | $\mathbf{6}$ | 559,000 | 276,000 | 304,000 | 445,000 | 321,000 | 187,000 | 137,000 |
|  | $\mathbf{7}$ | 466,000 | 266,000 | 287,000 | 262,000 | 257,000 | 201,000 | 105,000 |
|  | $\mathbf{8}$ | 634,000 | 272,000 | 286,000 | 193,000 | 174,000 | 198,000 | 123,000 |
|  | $\mathbf{9}$ | 578,000 | 284,000 | 225,000 | 154,000 | 93,000 | 174,000 | 103,000 |
|  | $\mathbf{1 0 +}$ | $1,460,000$ | 673,000 | 334,000 | 255,000 | 259,000 | 398,000 | 195,000 |


| Age |  | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{1}$ | 865,000 | $1,611,000$ | 266,686 | 407,730 | 263,184 | 306,951 | 296,100 | $1,893,453$ |
|  | $\mathbf{2}$ | 718,000 | 703,000 | $1,024,468$ | 653,838 | 305,180 | 107,935 | 353,949 | 534,221 |
|  | $\mathbf{3}$ | $1,340,000$ | 672,000 | 513,959 | $1,641,714$ | 621,085 | 367,962 | 421,560 | 632,361 |
|  | $\mathbf{4}$ | 791,000 | 753,000 | 301,627 | 569,094 | $1,571,236$ | 389,264 | 465,358 | 537,280 |
|  | $\mathbf{5}$ | 837,000 | 520,000 | 363,204 | 217,386 | 411,367 | $1,221,919$ | 615,994 | 323,324 |
|  | $\mathbf{6}$ | 708,000 | 577,000 | 258,038 | 154,044 | 191,241 | 281,120 | 800,201 | 497,458 |
|  | $\mathbf{7}$ | 139,000 | 299,000 | 159,153 | 109,580 | 107,005 | 174,256 | 253,818 | 663,133 |
|  | $\mathbf{8}$ | 50,000 | 78,000 | 49,431 | 79,663 | 64,769 | 90,429 | 159,797 | 232,420 |
| $\mathbf{9}$ | 25,000 | 27,000 | 5,060 | 31,987 | 38,118 | 79,014 | 59,670 | 98,415 |  |
| $\mathbf{1 0 +}$ | 38,000 | 95,000 | 9,570 | 11,706 | 17,476 | 30,614 | 41,811 | 82,521 |  |


| Age |  | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{1}$ | $2,131,494$ | $1,656,926$ | 788,200 | $1,814,851$ | $4,363,690$ | $1,821,053$ | $3,742,841$ | $2,140,731$ |
|  | $\mathbf{2}$ | $1,519,327$ | $4,181,175$ | $1,549,100$ | $1,192,657$ | $4,486,315$ | $3,232,244$ | $4,073,497$ | $4,394,444$ |
|  | $\mathbf{3}$ | 904,074 | $3,541,231$ | $5,820,800$ | $3,465,739$ | $2,962,163$ | $3,291,844$ | $8,378,955$ | $6,675,323$ |
|  | $\mathbf{4}$ | 577,676 | $1,044,897$ | $3,460,600$ | $5,014,862$ | $3,806,520$ | $2,242,722$ | $4,824,590$ | $6,649,684$ |
|  | $\mathbf{5}$ | 295,671 | 383,658 | 412,800 | $1,550,063$ | $2,592,933$ | $1,824,047$ | $2,035,096$ | $3,023,013$ |
|  | $\mathbf{6}$ | 251,642 | 322,777 | 207,200 | 513,663 | 585,666 | $1,647,122$ | $1,117,179$ | $1,267,219$ |
|  | $\mathbf{7}$ | 282,056 | 303,058 | 151,200 | 213,057 | 170,020 | 344,403 | 400,022 | 645,205 |
|  | $\mathbf{8}$ | 406,910 | 264,105 | 153,100 | 151,429 | 97,032 | 168,848 | 121,280 | 247,303 |
| $\mathbf{9}$ | 104,320 | 212,452 | 68,800 | 58,277 | 76,624 | 102,576 | 19,701 | 74,872 |  |
| $\mathbf{1 0 +}$ | 169,235 | 85,513 | 140,500 | 139,791 | 66,410 | 142,743 | 27,493 | 36,540 |  |

[^2]Table 4.5.5.1. Blue Whiting: Mean weights-at-age in the total catch and stock in 1981-2004.

| Age |  | $\mathbf{1 9 8 1}$ | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{0}$ | 0.038 | 0.018 | 0.020 | 0.026 | 0.016 | 0.030 | 0.023 |
|  | $\mathbf{1}$ | 0.052 | 0.045 | 0.046 | 0.035 | 0.038 | 0.040 | 0.048 |
|  | $\mathbf{2}$ | 0.065 | 0.072 | 0.074 | 0.078 | 0.074 | 0.073 | 0.086 |
|  | $\mathbf{3}$ | 0.103 | 0.111 | 0.118 | 0.089 | 0.097 | 0.108 | 0.106 |
|  | $\mathbf{4}$ | 0.125 | 0.143 | 0.140 | 0.132 | 0.114 | 0.130 | 0.124 |
|  | $\mathbf{5}$ | 0.141 | 0.156 | 0.153 | 0.153 | 0.157 | 0.165 | 0.147 |
|  | $\mathbf{6}$ | 0.155 | 0.177 | 0.176 | 0.161 | 0.177 | 0.199 | 0.177 |
|  | $\mathbf{7}$ | 0.170 | 0.195 | 0.195 | 0.175 | 0.199 | 0.209 | 0.208 |
|  | $\mathbf{8}$ | 0.178 | 0.200 | 0.200 | 0.189 | 0.208 | 0.243 | 0.221 |
|  | $\mathbf{9}$ | 0.187 | 0.204 | 0.204 | 0.186 | 0.218 | 0.246 | 0.222 |
|  | $\mathbf{1 0 +}$ | 0.213 | 0.231 | 0.228 | 0.206 | 0.237 | 0.257 | 0.254 |


| Age |  | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{0}$ | 0.014 | 0.034 | 0.036 | 0.024 | 0.028 | 0.033 | 0.022 |
|  | $\mathbf{1}$ | 0.059 | 0.045 | 0.055 | 0.057 | 0.066 | 0.061 | 0.064 |
|  | $\mathbf{2}$ | 0.079 | 0.070 | 0.091 | 0.083 | 0.082 | 0.087 | 0.091 |
|  | $\mathbf{3}$ | 0.103 | 0.106 | 0.107 | 0.119 | 0.109 | 0.108 | 0.118 |
|  | $\mathbf{4}$ | 0.126 | 0.123 | 0.136 | 0.140 | 0.137 | 0.137 | 0.143 |
|  | $\mathbf{5}$ | 0.148 | 0.147 | 0.174 | 0.167 | 0.163 | 0.164 | 0.154 |
|  | $\mathbf{6}$ | 0.158 | 0.168 | 0.190 | 0.193 | 0.177 | 0.189 | 0.167 |
|  | $\mathbf{7}$ | 0.171 | 0.175 | 0.206 | 0.226 | 0.200 | 0.207 | 0.203 |
|  | $\mathbf{8}$ | 0.203 | 0.214 | 0.230 | 0.235 | 0.217 | 0.217 | 0.214 |
|  | $\mathbf{9}$ | 0.224 | 0.217 | 0.232 | 0.284 | 0.225 | 0.247 | 0.206 |
|  | $\mathbf{1 0 +}$ | 0.253 | 0.256 | 0.266 | 0.294 | 0.281 | 0.254 | 0.256 |


| Age |  | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{0}$ | 0.031 | 0.033 | 0.035 | 0.031 | 0.038 | 0.021 | 0.019 |
|  | $\mathbf{1}$ | 0.047 | 0.048 | 0.063 | 0.057 | 0.050 | 0.054 | 0.049 |
|  | $\mathbf{2}$ | 0.072 | 0.072 | 0.078 | 0.075 | 0.078 | 0.074 | 0.075 |
|  | $\mathbf{3}$ | 0.102 | 0.094 | 0.088 | 0.086 | 0.094 | 0.093 | 0.098 |
|  | $\mathbf{4}$ | 0.121 | 0.125 | 0.109 | 0.104 | 0.108 | 0.115 | 0.108 |
|  | $\mathbf{5}$ | 0.140 | 0.149 | 0.142 | 0.133 | 0.129 | 0.132 | 0.131 |
|  | $\mathbf{6}$ | 0.166 | 0.178 | 0.170 | 0.156 | 0.163 | 0.155 | 0.148 |
|  | $\mathbf{7}$ | 0.177 | 0.183 | 0.199 | 0.179 | 0.186 | 0.173 | 0.168 |
|  | $\mathbf{8}$ | 0.183 | 0.188 | 0.193 | 0.187 | 0.193 | 0.233 | 0.146 |
|  | $\mathbf{9}$ | 0.203 | 0.221 | 0.192 | 0.232 | 0.231 | 0.224 | 0.232 |
|  | $\mathbf{1 0 +}$ | 0.232 | 0.248 | 0.245 | 0.241 | 0.243 | 0.262 | 0.258 |

Table 4.5.6.1 Blue whiting. Length at age composition (cm) of the landings from the main fisheries in 200

| Age | Norwegian Sea fishery <br> (Subareas I, II and <br> Divisions Va, XIVa,b) | Spawning area fishery <br> (Divisions Vb, Vla,b, <br> VIIb,c) | Directed and mixed <br> fisheries (Division IIIa <br> and Subarea IV) | Southern areas fishery <br> (Subareas VIII and IX, <br> and Divisions VIId, e,g-k) |
| ---: | ---: | ---: | ---: | ---: |
| 0 | 13.4 |  | 16.3 | 19.5 |
| 1 | 20.0 | 20.6 | 18.6 | 20.4 |
| 2 | 22.7 | 23.6 | 22.1 | 22.4 |
| 3 | 25.1 | 25.7 | 24.9 | 25.3 |
| 4 | 26.4 | 26.9 | 27.3 | 27.1 |
| 5 | 28.0 | 28.9 | 29.1 | 31.2 |
| 6 | 29.5 | 30.5 | 29.6 | 34.3 |
| 6 | 30.2 | 31.4 | 29.7 | 35.6 |
| 7 | 31.3 | 32.0 | 35.5 | 37.3 |
| 8 | 33.3 | 34.1 | 36.0 | 34.5 |
| 9 | 37.8 | 38.5 | 40.9 |  |
| $10+$ |  |  |  |  |

Table 4.5.7.1. Blue Whiting: natural mortality and proportion of maturation at age. Natural mortality is assumed to be the same in all years. The values for the maturity-ogive were estimated by the 1994 WG (ICES 1995/Assess:7).

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Proportion of mature | 0.00 | 0.11 | 0.40 | 0.82 | 0.86 | 0.91 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |
| Natural mortality | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |

Table 4.5.9.1.1. Blue whiting. Age stratified Spanish cpue (not used in the assessment).

| Numbers | age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | total |
| 1982 |  |  |  |  |  |  |  |  |
| 1983 |  | 7196 | 16392 | 9311 | 7476 | 6326 | 1718 | 48419 |
| 1984 |  | 13710 | 27286 | 14845 | 4836 | 1755 | 1750 | 64182 |
| 1985 |  | 14573 | 23823 | 14126 | 6256 | 1232 | 217 | 60227 |
| 1986 |  | 3721 | 14131 | 14745 | 7113 | 1278 | 505 | 41493 |
| 1987 |  | 25328 | 13153 | 6664 | 2938 | 1029 | 166 | 49278 |
| 1988 |  | 7778 | 21473 | 18436 | 6391 | 1300 | 781 | 56159 |
| 1989 |  | 15272 | 18486 | 17160 | 8374 | 3760 | 1003 | 64055 |
| 1990 |  | 21444 | 19407 | 5194 | 1803 | 1357 | 451 | 49656 |
| 1991 |  | 15924 | 15370 | 4989 | 2329 | 1045 | 440 | 40097 |
| 1992 |  | 10007 | 24235 | 9671 | 4316 | 1194 | 462 | 49885 |
| 1993 |  | 4036 | 13991 | 22493 | 7979 | 1354 | 658 | 50511 |
| 1994 |  | 543 | 6066 | 15917 | 7474 | 2990 | 1055 | 34045 |
| 1995 |  | 9090 | 14409 | 6833 | 4551 | 1990 | 623 | 37496 |
| 1996 |  | 3905 | 14557 | 14449 | 3931 | 3639 | 1834 | 42315 |
| 1997 |  | 8742 | 15875 | 11134 | 3698 | 1046 | 450 | 40945 |
| 1998 |  | 5884 | 13236 | 9803 | 10844 | 5229 | 1153 | 46149 |
| 1999 |  | 2048 | 10268 | 20242 | 9833 | 6287 | 3047 | 51725 |
| 2000 |  | 6207 | 15518 | 13987 | 5375 | 1264 | 1414 | 43765 |
| 2001 |  | 16223 | 16488 | 6830 | 1620 | 1148 | 162 | 42471 |
| 2002 |  | 10520 | 13725 | 10265 | 3385 | 336 | 69 | 38300 |
| 2003 |  | 9069 | 10461 | 6517 | 3983 | 1932 | 737 | 32699 |

Table 4.6.1.1.1 Blue Whiting: Age and length distribution of blue whiting in the international surveys by R/Vs "Atlantniro", "Celtic Explorer", "Fridtjof Nansen", "G.
O. Sars", "Magnus Heinason", and "Tridens"

| Length(cm) | Age in years (year class) |  |  |  |  |  |  |  |  |  |  | Num- <br> bers <br> $\left(10^{6}\right)$ | Bio- <br> mass <br> $\left(10^{6}\right.$ <br> kg ) | Mean weight (g) | Prop. mature * (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{array}{r}1 \\ 2004 \\ \hline\end{array}$ | 2 2003 | 3 2002 | $\begin{array}{r}4 \\ 2001 \\ \hline\end{array}$ | $\begin{array}{r}5 \\ 2000 \\ \hline\end{array}$ | $\begin{array}{r}6 \\ 1999 \\ \hline\end{array}$ | $\begin{array}{r}7 \\ 1998 \\ \hline\end{array}$ | $\begin{array}{r}8 \\ 1997 \\ \hline\end{array}$ |  | $\begin{array}{r} 10 \\ 1995 \\ \hline \end{array}$ | 111 |  |  |  |  |
| 13.0-14.0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 13.1 | 8 |
| 14.0-15.0 | 46 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 46 | 0.6 | 14.1 | 7 |
| 15.0-16.0 | 451 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 485 | 9 | 17.8 | 11 |
| 16.0-17.0 | 985 | 28 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1013 | 21 | 21.1 | 13 |
| 17.0-18.0 | 861 | 90 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 952 | 24 | 25.4 | 17 |
| 18.0-19.0 | 756 | 91 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 847 | 26 | 30.7 | 21 |
| 19.0-20.0 | 272 | 541 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 813 | 30 | 37.4 | 54 |
| 20.0-21.0 | 119 | 1125 | 25 | 0 | 0 | 10 | 0 | 0 | 0 | 0 | 0 | 1279 | 52 | 40.4 | 79 |
| 21.0-22.0 | 36 | 703 | 395 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1134 | 54 | 47.2 | 85 |
| 22.0-23.0 | 33 | 419 | 1342 | 148 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1941 | 111 | 57.2 | 85 |
| 23.0-24.0 | 0 | 823 | 3034 | 620 | 199 | 0 | 0 | 0 | 0 | 0 | 0 | 4676 | 294 | 62.9 | 86 |
| 24.0-25.0 | 49 | 262 | 4526 | 3507 | 1891 | 0 | 0 | 0 | 0 | 0 | 0 | 10236 | 711 | 69.5 | 91 |
| 25.0-26.0 | 0 | 204 | 5243 | 6608 | 3628 | 472 | 0 | 0 | 0 | 0 | 0 | 16155 | 1246 | 77.1 | 95 |
| 26.0-27.0 | 20 | 0 | 2645 | 6827 | 6516 | 579 | 16 | 0 | 0 | 0 | 0 | 16603 | 1404 | 84.6 | 97 |
| 27.0-28.0 | 0 | 0 | 1240 | 4270 | 5719 | 759 | 71 | 5 | 0 | 0 | 0 | 12063 | 1140 | 94.5 | 98 |
| 28.0-29.0 | 0 | 0 | 235 | 2348 | 3352 | 1282 | 254 | 85 | 0 | 0 | 0 | 7555 | 805 | 107 | 99 |
| 29.0-30.0 | 0 | 0 | 74 | 908 | 3285 | 1095 | 249 | 24 | 0 | 0 | 0 | 5635 | 663 | 118 | 99 |
| 30.0-31.0 | 0 | 0 | 9 | 238 | 1177 | 1484 | 68 | 129 | 37 | 0 | 0 | 3143 | 419 | 133 | 100 |
| 31.0-32.0 | 0 | 0 | 8 | 19 | 833 | 1480 | 311 | 18 | 5 | 0 | 0 | 2673 | 397 | 148 | 100 |
| $32.0-33.0$ | 0 | 0 | 0 | 86 | 11 | 601 | 302 | 62 | 47 | 0 | 0 | 1108 | 183 | 165 | 100 |
| 33.0-34.0 | 0 | 0 | 0 | 0 | 11 | 347 | 295 | 146 | 0 | 0 | 0 | 799 | 146 | 183 | 100 |
| $34.0-35.0$ | 0 | 0 | 0 | 0 | 0 | 142 | 295 | 81 | 79 | 2 | 4 | 602 | 121 | 201 | 100 |
| 35.0-36.0 | 0 | 0 | 0 | 0 | 7 | 9 | 61 | 8 | 43 | 0 | 0 | 128 | 27 | 209 | 100 |
| $36.0-37.0$ | 0 | 0 | 0 | 0 | 31 | 37 | 47 | 140 | 0 | 0 | 0 | 254 | 63 | 247 | 100 |
| 37.0-38.0 | 0 | 0 | 0 | 0 | 0 | 0 | 44 | 3 | 14 | 0 | 0 | 62 | 15 | 241 | 100 |
| 38.0-39.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 13 | 14 | 0 | 0 | 28 | 8 | 282 | 100 |
| 39.0-40.0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 10 | 43 | 0 | 0 | 58 | 18 | 311 | 100 |
| 40.0-41.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 39 | 0 | 0 | 45 | 17 | 382 | 100 |
| 41.0-42.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 2 | 0.5 | 343 | 100 |
| TSN ( $10^{6}$ ) | 3631 | 4320 | 18774 | 25579 | 26660 | 8298 | 2016 | 728 | 323 | 2 | 4 | 90336 |  |  |  |
| TSB ( $10^{6} \mathrm{~kg}$ ) | 99 | 217 | 1377 | 2194 | 2546 | 1046 | 320 | 128 | 76 | 0.5 | 0.7 | 8005 |  |  |  |
| Avg. length (cm) | 17.6 | 21.6 | 25.0 | 26.4 | 27.4 | 29.8 | 31.9 | 33.0 | 35.6 | 34.9 | 34.5 | 26.3 |  |  |  |
| Avg. weight (g) | 27.3 | 50.2 | 73.3 | 85.8 | 95.5 | 126 | 159 | 176 | 236 | 212 | 183 | 88.6 |  |  |  |
| Cond. (g/dm ${ }^{3}$ ) | 5.0 | 5.0 | 4.7 | 4.7 | 4.6 | 4.8 | 4.9 | 4.9 | 5.2 | 5.0 | 4.5 | 4.9 |  |  |  |
| \% mature* | 13 | 79 | 93 | 93 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 92 |  |  |  |
| \% of SSB | 0 | 2 | 17 | 27 | 33 | 14 | 4 | 2 | 1 | 0 | 0 |  |  |  |  |

Table 4.6.1.2.1. Age and length distribution of blue whiting in the survey by R/V "G. O.
Sars" west of the British Isles, March-April 2005.

| Length <br> (cm) | $\begin{array}{r} 1 \\ 2004 \\ \hline \end{array}$ | $\begin{array}{r} 2 \\ 2003 \\ \hline \end{array}$ | $\begin{array}{r} 3 \\ 2002 \\ \hline \end{array}$ | Age in <br> 4 <br> 2001 | years 5 2000 | (year c 6 1999 | class) 7 1998 | $\begin{array}{r} 8 \\ 1997 \\ \hline \end{array}$ | $\begin{array}{r} 9 \\ 1996 \\ \hline \end{array}$ | $\begin{array}{r} 10 \\ 1995 \\ \hline \end{array}$ | $\begin{aligned} & \text { Num- } \\ & \text { bers } \\ & \left(10^{6}\right) \\ & \hline \end{aligned}$ | $\left\|\begin{array}{c}\text { Bio- } \\ \text { mass } \\ \left(10^{6} \mathrm{~kg}\right)\end{array}\right\|$ | Mean weight (g) | $\begin{array}{\|c\|} \hline \text { Ma- } \\ \text { ture } \\ (\%) \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 14.0-15.0 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 16 | 0.2 | 14 | 0 |
| 15.0-16.0 | 48 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 48 | 0.8 | 16.8 | 0 |
| 16.0-17.0 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 41 | 0.8 | 19.0 | 0 |
| 17.0-18.0 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 11 | 0.3 | 24.2 | 0 |
| 18.0-19.0 | 98 | 26 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 124 | 3.4 | 27.1 | 0 |
| 19.0-20.0 | 95 | 109 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 203 | 6.4 | 31.6 | 11 |
| 20.0-21.0 | 23 | 602 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 624 | 21 | 34.1 | 14 |
| 21.0-22.0 | 23 | 78 | 103 | 0 | 0 | 0 | 0 | 0 | 0 |  | 204 | 9.4 | 46.0 | 54 |
| 22.0-23.0 | 15 | 0 | 534 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 549 | 28 | 50.7 | 94 |
| 23.0-24.0 | 0 | 409 | 2822 | 384 | 13 | 0 | 0 | 0 | 0 | 0 | 3627 | 215 | 59.3 | 98 |
| 24.0-25.0 | 0 | 124 | 4918 | 2439 | 1876 | 0 | 0 | 0 | 0 | 0 | 9356 | 618 | 66.1 | 99 |
| 25.0-26.0 | 0 | 110 | 6309 | 7595 | 3801 | 734 | 0 | 0 | 0 |  | 18549 | 1385 | 74.7 | 100 |
| 26.0-27.0 | 0 | 0 | 2764 | 8872 | 8126 | 344 | 0 | 0 | 0 | 0 | 20105 | 1664 | 82.8 | 100 |
| 27.0-28.0 | 0 | 0 | 1798 | 5806 | 6301 | 976 | 26 | 0 | 0 | 0 | 14907 | 1384 | 92.8 | 100 |
| 28.0-29.0 | 0 | 0 | 571 | 3367 | 4246 | 1254 | 425 | 165 | 0 | 0 | 10028 | 1052 | 105 | 100 |
| 29.0-30.0 | 0 | 0 | 150 | 1657 | 4325 | 1560 | 372 | 36 | 0 | 0 | 8100 | 938 | 116 | 100 |
| 30.0-31.0 | 0 | 0 | 0 | 206 | 1613 | 907 | 37 | 343 | 0 | 0 | 3106 | 401 | 129 | 100 |
| 31.0-32.0 | 0 | 0 | 0 | 29 | 1318 | 979 | 286 | 16 | 29 | 0 | 2658 | 384 | 144 | 100 |
| 32.0-33.0 | 0 | 0 | 0 | 103 | 39 | 233 | 413 | 0 | 25 | 0 | 813 | 131 | 162 | 100 |
| 33.0-34.0 | 0 | 0 | 0 | 0 | 37 | 230 | 113 | 132 | 0 | 0 | 512 | 92 | 179 | 100 |
| 34.0-35.0 | 0 | 0 | 0 | 0 | 0 | 16 | 295 | 2 | 189 | 15 | 516 | 99 | 191 | 100 |
| 35.0-36.0 | 0 | 0 | 0 | 0 | 0 | 12 | 3 | 0 | 91 | 0 | 106 | 23 | 216 | 100 |
| 36.0-37.0 | 0 | 0 | 0 | 0 | 14 | 175 | 14 | 13 | 0 | 0 | 215 | 48 | 221 | 100 |
| 37.0-38.0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
| 38.0-39.0 | 0 | 0 |  | 0 | 0 | 27 | 7 | 38 | 0 | 0 | 71 | 20 | 285 | 100 |
| 39.0-40.0 | 0 | 0 | 0 | 0 | 0 |  | 2 | 2 | 0 | 0 | 12 | 3.1 | 263 | 100 |
| 40.0-41.0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0.4 | 410 | 100 |
| TSN ( $10^{6}$ ) | 370 | 14561 | 9968 | 304593 | 31708 | 7455 | 1993 | 747 | 333 | 15 | 94503 |  |  |  |
| TSB ( $10^{6} \mathrm{~kg}$ ) | 11 | 69 | 1469 | 2608 | 3025 | 882 | 287 | 107 | 64 | 2.9 | 8527 |  |  |  |
| Avg. length (cm) | 18.4 | 22.0 | 25.3 | 26.7 | 27.6 | 29.4 | 31.3 | 31.1 | 34.4 | 34.5 | 27.0 |  |  |  |
| Avg. weight (g) | 29.4 | 47.5 | 73.6 | 85.6 | 95.4 | 118 | 144 | 143 | 194 | 192 | 90.2 |  |  |  |
| Condition | 4.7 | 4.5 | 4.5 | 4.5 | 4.5 | 4.7 | 4.7 | 4.8 | 4.8 | 4.7 | 4.6 |  |  |  |
| \% mature | 10 | 50 | 99 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 99.5 |  |  |  |
| \% of SSB | 0 | 0 | 17 | 31 | 36 | 10 | 3 | 1 | 1 | 0 |  |  |  |  |

Table 4.6.1.2.2. Stock estimates of blue whiting in the Norwegian spawning stock surveys west of the British Isles.

| Year | Abundance, $10^{9}$ individuals <br> total | Biomass, mill. tonnes <br> spawning <br> total | Mean weight, <br> spawning | Mean length, <br> g | cm |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 63 | 56 | 6.3 | 5.7 | 101 | 27.1 |
| 1991 | 42 | 41 | 5.1 | 4.8 | 116 | 27.8 |
| 1992 | 38 | 37 | 4.3 | 4.2 | 111 | 27.5 |
| 1993 | 42 | 40 | 5.2 | 5.0 | 125 | 28.6 |
| 1994 | 27 | 26 | 4.1 | 4.1 | 153 | 31.1 |
| 1995 | 62 | 45 | 6.7 | 6.1 | 108 | 26.9 |
| 1996 | 52 | 36 | 5.1 | 4.5 | 94.9 | 25.5 |
| 1997 |  |  |  |  |  |  |
| 1998 | 80 | 57 | 5.5 | 4.7 | 68.3 | 23.2 |
| 1999 | 120 | 110 | 8.9 | 8.5 | 74.4 | 25.0 |
| 2000 | 102 | 90 | 8.3 | 7.8 | 80.7 | 25.5 |
| 2001 | 97 | 72 | 6.7 | 5.6 | 69.0 | 24.1 |
| 2002 | 176 | 147 | 12.2 | 10.9 | 69.3 | 24.2 |
| 2003 | 160 | 132 | 11.4 | 10.4 | 71.6 | 24.6 |
| 2004 | 137 | 128 | 11.4 | 10.9 | 83.2 | 26.1 |
| 2005 | 95 | 93 | 8.5 | 8.5 | 90.2 | 27.0 |

Table 4.6.1.3.1 Age stratified acoustic survey estimates of blue whiting in the spawning area by Russian vessels Number in millions, biomass in thousand t , mean length in cm , mean weight in grams

| Numbers | age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | total |
| 1981 |  |  |  |  |  |  |  |  |  |  |  |
| 1982 |  |  | 540 | 2750 | 1340 | 1380 | 1570 | 2350 | 1730 | 1290 | 12950 |
| 1983 |  |  | 2330 | 2930 | 9390 | 3880 | 1970 | 1370 | 780 | 660 | 23310 |
| 1984 |  |  | 2900 | 800 | 1100 | 4200 | 2200 | 1200 | 1700 | 1200 | 15300 |
| 1985 |  |  | 13220 | 930 | 580 | 1780 | 860 | 610 | 580 | 540 | 19100 |
| 1986 |  |  | 18750 | 23180 | 2540 | 610 | 620 | 750 | 640 | 710 | 47800 |
| 1987 |  |  | 4480 | 19170 | 5860 | 1070 | 500 | 810 | 860 | 670 | 33420 |
| 1988 |  |  | 3710 | 4550 | 8610 | 4130 | 1270 | 480 | 250 | 260 | 23260 |
| 1989 |  |  | 11910 | 7120 | 6670 | 6970 | 4580 | 2750 | 1880 | 810 | 42690 |
| 1990 |  |  | 9740 | 12140 | 5740 | 2580 | 1470 | 220 | 80 | 10 | 31980 |
| 1991 |  |  | 10300 | 5350 | 5130 | 2630 | 1770 | 870 | 300 | 220 | 26570 |
| 1992 |  |  | 20010 | 6700 | 1350 | 440 | 390 | 170 | 0 | 0 | 29060 |
| 1993 |  |  | 4728 | 12337 | 5304 | 2249 | 1316 | 621 | 386 | 150 | 27091 |
| 1994 |  |  | no survey |  |  |  |  |  |  |  |  |
| 1995 |  |  | 12657 | 10028 | 8942 | 2651 | 1093 | 408 | 131 | 14 | 35924 |
| 1996 |  |  | 15285 | 10629 | 4897 | 6940 | 1482 | 653 | 85 | 0 | 39971 |
| 1997 |  |  | no survey |  |  |  |  |  |  |  |  |
| 1998 |  |  | no survey |  |  |  |  |  |  |  |  |
| 1999 |  |  | no survey |  |  |  |  |  |  |  |  |
| 2000 |  |  | no survey |  |  |  |  |  |  |  |  |
| 2001 |  |  | no comparab | survey |  |  |  |  |  |  |  |
| 2002 |  |  | no comparab | survey |  |  |  |  |  |  |  |
| 2003 |  |  | no comparab | survey |  |  |  |  |  |  |  |
| 2004 |  |  | no comparab | survey |  |  |  |  |  |  |  |

$\square$ used in the assessment

Table 4.6.1.4.1. Blue Whiting: Age- and length-stratified abundance estimate of blue whiting in the Norwegian Sea in May-June 2005. Data from R/Vs "Dana", "Magnus Heinason", "Arni Fridriksson", "Johan Hjort" and "G. O. Sars". Target strength used for blue whiting: $21.8 \log \mathrm{~L}-72.8 \mathrm{~dB}$.


Table 4.6.1.4.2. Blue whiting: Estimated stock biomass, numbers, length and weight at age for blue whiting in the standard survey area (between $8^{\circ} \mathrm{W}-20^{\circ} \mathrm{E}$ and north of $63^{\circ} \mathrm{N}$ ) in the international surveys 2000-2005.

| Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Numbers ( $10^{6}$ ) |  |  |  |  |  |  |  |  |  |  |  |  |
| 2000 | 48927 | 3133 | 3580 | 1668 | 201 | 5 |  |  |  |  |  | 57514 |
| 2001 | 85772 | 25110 | 7533 | 3020 | 2066 |  |  |  |  |  |  | 123501 |
| 2002 | 15251 | 46656 | 14672 | 4357 | 513 | 445 |  | 15 |  | 6 |  | 81915 |
| 2003 | 35688 | 21487 | 35372 | 4354 | 639 | 201 | 43 | 3 |  |  |  | 97787 |
| 2004 | 49254 | 22086 | 13292 | 8290 | 1495 | 533 | 83 | 39 |  |  |  | 95072 |
| 2005 | 54660 | 19904 | 13828 | 4714 | 1886 | 326 | 103 | 43 | 8 | 3 | 11 | 95486 |
| Biomass ( $10^{6} \mathrm{~kg}$ ) |  |  |  |  |  |  |  |  |  |  |  |  |
| 2000 | 1795 | 260 | 335 | 193 | 25 | 1 |  |  |  |  |  | 2608 |
| 2001 | 2735 | 1776 | 763 | 418 | 322 |  |  |  |  |  |  | 6014 |
| 2002 | 651 | 2640 | 1289 | 526 | 76 | 64 |  | 3 |  | 2 |  | 5250 |
| 2003 | 1475 | 1539 | 2897 | 497 | 88 | 31 | 11 | 1 |  |  |  | 6538 |
| 2004 | 1643 | 1437 | 1188 | 886 | 193 | 77 | 13 | 6 |  |  |  | 5442 |
| 2005 | 1558 | 1204 | 1124 | 502 | 233 | 49 | 16 | 8 | 2 | 1 | 2 | 4699 |
| Length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |
| 2000 | 19.2 | 24.7 | 25.6 | 27.3 | 27.7 | 33.2 |  |  |  |  |  | 20.2 |
| 2001 | 18.2 | 23.4 | 26.3 | 28.8 | 29.8 |  |  |  |  |  |  | 20.2 |
| 2002 | 20.1 | 21.9 | 25.1 | 27.9 | 30.1 | 30.2 |  | 34.5 |  | 37.5 |  | 22.5 |
| 2003 | 20.1 | 23.5 | 24.5 | 27 | 28.9 | 29.9 | 34.5 | 33.5 |  |  |  | 22.8 |
| 2004 | 18.7 | 22.5 | 24.8 | 26.5 | 28.6 | 30.1 | 31.4 | 30.9 |  |  |  | 21.4 |
| 2005 | 17.9 | 22.3 | 24.3 | 26.5 | 28 | 30.3 | 31 | 32.7 | 32.7 | 30.5 | 38.5 | 20.4 |
| Weight (g) |  |  |  |  |  |  |  |  |  |  |  |  |
| 2000 | 36.7 | 83 | 93.5 | 116 | 122 | 225 |  |  |  |  |  | 45.3 |
| 2001 | 31.9 | 70.7 | 101 | 138 | 156 |  |  |  |  |  |  | 48.7 |
| 2002 | 42.7 | 56.6 | 87.8 | 121 | 147 | 145 |  | 210 |  | 269 |  | 64.1 |
| 2003 | 41.3 | 71.6 | 81.9 | 114 | 138 | 153 | 256 | 219 |  |  |  | 66.9 |
| 2004 | 33.4 | 65 | 89.4 | 107 | 129 | 144 | 162 | 160 |  |  |  | 57.2 |
| 2005 | 28.8 | 61.7 | 82.7 | 108 | 126 | 155 | 164 | 197.3 | 189.5 | 157.7 | 222 | 49.9 |

Table 4.6.1.5.1 Blue whiting. Age stratified acoustic survey estimates in the Icelandic EEZ in July (not used in the assessment)

| Year | Age (yrs) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ | Total |
| Numbers (millions) |  |  |  |  |  |  |  |  |  |  |  |
| 1999 | 14869 | 2100 | 1357 | 1772 | 5790 | 1344 | 316 | 50 | 15 | 42 | 27655 |
| 2000 | 10683 | 8594 | 934 | 523 | 1218 | 468 | 106 | 25 | 1 | 1 | 22553 |
| 2001 | 27305 | 4090 | 5215 | 1657 | 1614 | 398 | 132 | 37 | 6 | 2 | 40456 |
| 2002 | 3815 | 10785 | 3107 | 1436 | 1724 | 1430 | 727 | 178 | 47 | 5 | 23254 |
| 2003 | 5011 | 9363 | 6054 | 7430 | 3888 | 1350 | 852 | 581 | 91 | 33 | 34653 |
| 2004 | 10437 | 989 | 3970 | 3983 | 4854 | 2048 | 817 | 507 | 157 | 33 | 27795 |
| 2003* | 5011 | 9158 | 4899 | 4645 | 1918 | 646 | 218 | 227 | 91 | 6 | 26819 |

* An estimate calculated for an area that is comparable with the area covered by surveys in 1999-2002

| Biomass (tonnes*10^-3) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 265 | 163 | 127 | 201 | 764 | 212 | 55 | 13 | 4 | 14 | 1818 |
| 2000 | 186 | 624 | 85 | 63 | 167 | 78 | 22 | 5 | 0 | 0 | 1230 |
| 2001 | 661 | 295 | 568 | 211 | 231 | 66 | 22 | 8 | 1 | 0 | 2063 |
| 2002 | 77 | 746 | 297 | 160 | 217 | 203 | 114 | 31 | 13 | 1 | 1859 |
| 2003 | 68 | 555 | 600 | 853 | 503 | 200 | 147 | 102 | 19 | 9 | 3055 |
| 2004 | 169 | 59 | 354 | 429 | 620 | 305 | 153 | 100 | 33 | 8 | 2229 |


| Mean length $(\mathrm{cm})$ |  |  |  |  |  |  |  |  |  |  |  |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1999 | 13.5 | 23.5 | 25.6 | 26.3 | 27.6 | 29.0 | 30.3 | 33.7 | 35.5 | 36.8 |  |
| 2000 | 13.5 | 22.3 | 24.1 | 26.1 | 27.6 | 29.4 | 32.3 | 32.0 | 36.0 | 28.0 | 18.7 |
| 2001 | 15.1 | 22.4 | 25.3 | 26.4 | 28.1 | 29.9 | 31.8 | 32.6 | 33.0 | 37.0 | 18.3 |
| 2002 | 14.8 | 22.8 | 24.9 | 26.2 | 27.6 | 29.3 | 30.6 | 30.9 | 36.9 | 35.0 | 23.1 |
| 2003 | 13.2 | 21.3 | 25.1 | 26.4 | 27.9 | 29.2 | 31.0 | 31.1 | 33.9 | 31.0 | 23.4 |
| 2004 | 13.5 | 22.3 | 24.1 | 26.1 | 27.6 | 29.4 | 32.3 | 32.0 | 36.0 | 28.0 | 18.0 |


| Mean weight (gr) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 17.8 | 77.4 | 93.8 | 113 | 132 | 158 | 174 | 254 | 248 | 307 | 66 |
| 2000 | 17.5 | 72.4 | 88.5 | 117 | 137 | 164 | 211 | 188 | 245 | 322 | 50 |
| 2001 | 24.0 | 72.0 | 109.0 | 127 | 143 | 165 | 190 | 218 | 194 | 234 | 51 |
| 2002 | 20.3 | 69.2 | 95.5 | 111 | 126 | 142 | 157 | 173 | 267 | 205 | 80 |
| 2003 | 13.6 | 59.3 | 99.1 | 115 | 129 | 148 | 172 | 175 | 212 | 257 | 88 |
| 2004 | 16 | 59 | 89 | 108 | 128 | 149 | 188 | 197 | 208 | 242 | 95 |

Table 4.6.1.6.1. Blue Whiting. 1-group index from the Islandic Groundfish survey

| year | age1 |
| :--- | :--- |
| 1996 | 6.055 |
| 1997 | 2.803 |
| 1998 | 0.903 |
| 1999 | 5.308 |
| 2000 | 9.139 |
| 2001 | 4.902 |
| 2002 | 12.243 |
| 2003 | 14.142 |
| 2004 | 7.423 |
| 2005 | 6.326 |

Table 4.6.1.7.1 Age stratified acoustic survey estimates of blue whiting in the Norwegian Sea in July-August. Numbers in millions.

| Year\Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | Total |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1981 | 182 | 728 | 4542 | 3874 | 2678 | 2834 | 2964 | 2756 | 2054 | 22612 |
| 1982 | 184 | 460 | 1242 | 4715 | 3611 | 3128 | 2323 | 1679 | 874 | 18216 |
| 1983 | 22356 | 396 | 468 | 756 | 1404 | 576 | 468 | 432 | 324 | 27180 |
| 1984 | 30380 | 13916 | 833 | 392 | 539 | 539 | 343 | 49 | 49 | 47040 |
| 1985 | 5969 | 23876 | 12502 | 658 | 423 | 188 | 235 | 141 | 376 | 44368 |
| 1986 | 2324 | 2380 | 7224 | 6944 | 1876 | 952 | 336 | 308 | 140 | 22484 |
| 1987 | 8204 | 4032 | 5180 | 5572 | 1204 | 224 | 168 | 56 | 84 | 24724 |
| 1988 | 4992 | 2880 | 2640 | 3480 | 912 | 120 | 96 | 24 | 48 | 15192 |
| 1989 | 1172 | 1125 | 812 | 379 | 410 | 212 | 22 | 32 |  | 4164 |
| 1990 | no survey |  |  |  |  |  |  |  |  |  |
| 1991 | no survey |  |  |  |  |  |  |  |  |  |
| 1992 | 792 | 1134 | 6939 | 766 | 247 | 172 | 90 | 11 | 18 | 10169 |
| 1993 | 830 | 125 | 1070 | 6392 | 1222 | 489 | 248 | 58 | 88 | 10522 |
| 1994 | no survey |  |  |  |  |  |  |  |  |  |
| 1995 | 6974 | 2811 | 1999 | 1209 | 1622 | 775 | 173 | 61 |  | 15624 |
| 1996 | 23464 | 1057 | 899 | 649 | 436 | 505 | 755 | 69 | 41 | 27875 |
| 1997 | 30227 | 25638 | 1524 | 779 | 300 | 407 | 260 | 137 | 123 | 59395 |
| 1998 | 24244 | 47815 | 16282 | 556 | 212 | 100 | 64 | 10 | 255 | 89538 |
| 1999 | 14367 | 9750 | 23701 | 9754 | 1733 | 466 | 79 | 48 | 91 | 59989 |
| 2000 | 25813 | 3298 | 2721 | 3078 | 23 | 46 | 6 |  |  | 34985 |
| 2001 | 61470 | 22051 | 7883 | 3225 | 1824 | 156 | 12 |  | 68 | 96689 |
| 2002 | no survey |  |  |  |  |  |  |  |  |  |

Table 4.6.1.9.1 Stratified mean catch ( $\mathrm{Kg} / \mathrm{haul}$ and Number/haul) and standard error of BLUE WHITING in bottom trawl surveys in Spanish waters (Divisions VIIIc and IXa north). All surveys in September-October

| Kg/haul Year | $30-100 \mathrm{~m}$ |  | 101-200 m |  | 201-500 m |  | TOTAL 30-500 m |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | Mean | SD | Mean | SD | Mean | SD |
| 1985 | 9.50 | 5.87 | 119.75 | 45.99 | 68.18 | 13.79 | 92.83 | 28.24 |
| 1986 | 9.74 | 7.13 | 45.41 | 12.37 | 29.54 | 8.70 | 36.93 | 7.95 |
| 1987 | - | - | - | - | - | - | - | - |
| 1988 | 2.90 | 2.59 | 154.12 | 38.69 | 183.07 | 141.94 | 143.30 | 45.84 |
| 1989 | 14.17 | 12.03 | 76.92 | 17.08 | 18.79 | 6.23 | 59.00 | 11.68 |
| 1990 | 6.25 | 3.29 | 52.54 | 9.00 | 18.80 | 4.99 | 43.60 | 6.60 |
| 1991 | 64.59 | 34.65 | 126.41 | 26.06 | 46.07 | 18.99 | 97.10 | 17.16 |
| 1992 | 6.37 | 2.59 | 44.12 | 6.64 | 29.50 | 6.16 | 34.60 | 4.23 |
| 1993 | 1.06 | 0.63 | 14.07 | 3.73 | 51.08 | 22.02 | 22.59 | 6.44 |
| 1994 | 8.04 | 5.28 | 37.18 | 8.45 | 25.42 | 5.27 | 29.70 | 5.19 |
| 1995 | 19.97 | 13.87 | 36.43 | 4.82 | 15.97 | 4.10 | 28.52 | 3.66 |
| 1996 | 7.27 | 3.95 | 49.23 | 7.19 | 92.54 | 17.76 | 54.52 | 6.36 |
| 1997 | 17.87 | 7.35 | 44.68 | 10.52 | 57.14 | 16.60 | 42.62 | 7.29 |
| 1998 | 14.13 | 4.17 | 42.78 | 8.13 | 78.88 | 22.01 | 47.14 | 7.58 |
| 1999 | 93.01 | 14.60 | 112.39 | 19.92 | 169.21 | 50.26 | 124.66 | 17.85 |
| 2000 | 62.39 | 12.00 | 91.99 | 14.75 | 58.72 | 24.94 | 76.19 | 10.61 |
| 2001 | 8.35 | 3.31 | 50.18 | 10.09 | 52.41 | 16.71 | 42.02 | 7.02 |
| 2002 | 31.40 | 5.02 | 69.00 | 13.41 | 36.75 | 12.07 | 51.80 | 7.64 |
| 2003 | 42.52 | 12.22 | 71.40 | 11.01 | 46.43 | 11.42 | 58.13 | 6.92 |
| 2004 | 2.80 | 2.11 | 14.05 | 7.79 | 59.51 | 21.41 | 24.76 | 7.31 |


| Number/haul Year | $30-100 \mathrm{~m}$ |  | 101-200 m |  | 201-500 m |  | TOTAL 30-500 m |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | Mean | SD | Mean | SD | Mean | SD |
| 1985 | 267 | 181.71 | 3669 | 1578.86 | 1377 | 262.98 | 2644 | 963.20 |
| 1986 | 368 | 237.56 | 2486 | 1006.67 | 752 | 238.87 | 1763 | 616.40 |
| 1987 | - | - | - | - | - | - | - | - |
| 1988 | 83 | 71.74 | 6112 | 1847.36 | 7276 | 6339.88 | 5694 | 2086.00 |
| 1989 | 629 | 537.29 | 3197 | 876.75 | 566 | 213.11 | 2412 | 599.00 |
| 1990 | 220 | 115.48 | 2219 | 426.46 | 578 | 185.43 | 1722 | 276.00 |
| 1991 | 2922 | 1645.73 | 5563 | 1184.69 | 1789 | 847.33 | 4214 | 780.88 |
| 1992 | 124 | 50.81 | 1412 | 233.99 | 845 | 199.12 | 1069 | 146.87 |
| 1993 | 14 | 8.61 | 257 | 69.61 | 894 | 427.77 | 401 | 124.53 |
| 1994 | 346 | 234.12 | 2002 | 456.50 | 997 | 245.91 | 1487 | 689.00 |
| 1995 | 1291 | 864.97 | 2004 | 341.48 | 485 | 137.81 | 1493 | 240.37 |
| 1996 | 147 | 82.71 | 1167 | 167.20 | 2097 | 385.23 | 1263 | 142.30 |
| 1997 | 552 | 235.60 | 1443 | 361.89 | 1183 | 323.14 | 1180 | 209.94 |
| 1998 | 351 | 105.96 | 1463 | 320.26 | 2012 | 590.04 | 1387 | 234.82 |
| 1999 | 2508 | 427.20 | 4388 | 849.80 | 6119 | 2026.40 | 4490 | 727.90 |
| 2000 | 2267 | 414.97 | 3930 | 604.11 | 2009 | 859.71 | 3027 | 400.87 |
| 2001 | 171 | 77.34 | 1310 | 263.84 | 1232 | 381.49 | 1048 | 172.74 |
| 2002 | 771 | 90.34 | 2526 | 499.30 | 1075 | 331.09 | 1739 | 268.70 |
| 2003 | 1320 | 384.25 | 2791 | 554.16 | 1513 | 454.02 | 2114 | 317.68 |
| 2004 | 31 | 22.77 | 336 | 154.33 | 1472 | 736.78 | 599 | 225.74 |


| Year | Month | 20-100 m |  | 100-200 m |  | 200-500 m |  | 500-750 m |  | TOTAL |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | y | sy | y | sy | y | sy | y | sy | y | sy |
| 1990 | July | 2 | 2 | 153 | 103 | 242 | 42 | 50 | 5 | 96 | 35 |
|  | October | 11 | 5 | 90 | 28 | 762 | 234 | 42 | 10 | 153 | 35 |
| 1991 | July | 1 | 1 | 140 | 40 | 268 | 38 | 64 | 18 | 98 | 15 |
|  | October | 8 | 5 | 83 | 18 | 259 | 53 | 121 | 27 | 91 | 11 |
| 1992 | February | 7 | 7 | 43 | 35 | 249 | 21 | 73 | 3 | 68 | 12 |
|  | July | 1 | 1 | 29 | 18 | 216 | 43 | 27 | 5 | 47 | 9 |
|  | October | 1 | 1 | 22 | 7 | 208 | 44 | 80 | 3 | 54 | 7 |
| 1993 | February | 0 | 0 | 19 | 14 | 105 | 31 | 36 | 0 | 42 | 10 |
|  | July | 0 | 0 | 3 | 3 | 151 | 28 | 55 | 5 | 34 | 4 |
|  | November | 0 | 0 | 90 | 0 | 189 | 43 | 6 | 1 | 86 | 9 |
| 1994 | October | 0 | 0 | 374 | 30 | 283 | 32 | 49 | 7 | 174 | 11 |
| 1995 | July | 0 | 0 | 18 | 14 | 130 | 20 | 52 | 3 | 35 | 5 |
|  | October | 18 | 15 | 103 | 21 | 328 | 91 | 31 | 12 | 94 | 16 |
| 1996 | October | 25 | 24 | 12 | 2 | 36 | 6 | 25 | 7 | 22 | 8 |
| 1997 | June | 0 | 0 | 3 | 3 | 116 | 42 | 45 | 12 | 27 | 7 |
|  | October | 2 | 1 | 54 | 20 | 77 | 13 | 7 | 2 | 32 | 8 |
| 1998 | July | 0 | 0 | 8 | 5 | 105 | 17 | 38 | 3 | 25 | 3 |
|  | October | 1 | 1 | 384 | 87 | 427 | 101 | 20 | 2 | 212 | 36 |
| 1999 | July | 1 | 0 | 60 | 21 | 66 | 19 | 25 | 2 | 37 | 9 |
|  | October | 0 | 0 | 69 | 16 | 80 | 20 | 18 | 8 | 41 | 7 |
| 2000 | July | 23 | 13 | 109 | 34 | 116 | 10 | 63 | 6 | 75 | 13 |
|  | October | 11 | 4 | 155 | 53 | 196 | 22 | 54 | 4 | 99 | 19 |
| 2001 | July | 18 | 7 | 238 | 37 | 305 | 116 | 57 | 14 | 152 | 23 |
|  | October | 106 | 6 | 474 | 224 | 294 | 66 |  | 0 | 295 | 97 |
| 2002 | October | 19 | 12 | 176 | 81 | 180 | 24 |  | 0 | 116 | 34 |
| 2003 | October | 24 | 10 | 114 | 14 | 119 | 30 | 34 | 6 | 76 | 8 |
| 2004 | October | 0 | 0 | 44 | 10 | 380 | 27 |  |  | 84 | 15 |

Table 4.6.1.11.1 Catch in number of 1-group (age 1) blue whiting from the spring (March) bottom trawl surveys on the Faroe plateau 1996-2005. The number of 1-group in 2004 was taken out by age readings while for other years they were taken out from a visual inspection of the length distributions. There was a clear separation of the 1 and 2-group in the data (see Table 6.4.2.3).

| Year | 1 -group |
| ---: | ---: |
| 1994 | 1388 |
| 1995 | 1171 |
| 1996 | 4442 |
| 1997 | 1239 |
| 1998 | 262 |
| 1999 | 1108 |
| 2000 | 782 |
| 2001 | 2058 |
| 2002 | 3885 |
| 2003 | 873 |
| 2004 | 13016 |
| 2005 | 22653 |

Table 4.6.1.11.2 Length distribution (cm) of blue whiting from the spring (March) bottom trawl surveys on the Faroe plateau 1994-2005. Shaded areas in the years 1994-2005 indicate 1-group fish separated from visual inspection of the length distributions (in lack of otolith samples from the catch). In 2004 the 1 -group was separated by age readings from the survey.

|  |  |  |  |  | Year |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Length | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 | 18 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 | 107 | 149 | 0 | 0 | 0 | 11 | 0 | 0 | 159 | 512 |
| 15 | 0 | 0 | 702 | 347 | 0 | 5 | 1 | 68 | 0 | 0 | 1285 | 2402 |
| 16 | 1 | 8 | 865 | 207 | 6 | 92 | 51 | 289 | 0 | 16 | 3508 | 5617 |
| 17 | 25 | 0 | 585 | 303 | 31 | 178 | 126 | 356 | 124 | 303 | 3106 | 5212 |
| 18 | 198 | 53 | 669 | 141 | 72 | 263 | 133 | 398 | 262 | 249 | 2035 | 4498 |
| 19 | 442 | 146 | 567 | 79 | 64 | 317 | 200 | 439 | 553 | 162 | 1932 | 2244 |
| 20 | 399 | 111 | 539 | 132 | 45 | 200 | 137 | 230 | 1050 | 46 | 1005 | 1302 |
| 21 | 271 | 446 | 250 | 130 | 44 | 53 | 88 | 267 | 843 | 97 | 589 | 866 |
| 22 | 46 | 341 | 140 | 263 | 98 | 222 | 46 | 1299 | 1053 | 56 | 1544 | 1097 |
| 23 | 6 | 66 | 164 | 331 | 157 | 386 | 105 | 2235 | 1120 | 377 | 1471 | 853 |
| 24 | 0 | 112 | 793 | 356 | 183 | 623 | 221 | 2348 | 1047 | 472 | 2401 | 1536 |
| 25 | 0 | 224 | 1051 | 139 | 179 | 1022 | 312 | 2664 | 1859 | 856 | 2462 | 2512 |
| 26 | 0 | 291 | 1456 | 94 | 136 | 1259 | 668 | 2687 | 2670 | 602 | 3909 | 3394 |
| 27 | 8 | 194 | 841 | 82 | 93 | 789 | 1049 | 2493 | 4125 | 691 | 3100 | 2140 |
| 28 | 4 | 89 | 548 | 61 | 57 | 529 | 1148 | 2954 | 4564 | 240 | 2659 | 1698 |
| 29 | 5 | 49 | 195 | 93 | 48 | 314 | 1105 | 1774 | 5014 | 184 | 2080 | 1530 |
| 30 | 23 | 27 | 159 | 16 | 28 | 105 | 939 | 731 | 4852 | 229 | 965 | 1862 |
| 31 | 36 | 28 | 108 | 38 | 43 | 140 | 549 | 526 | 2924 | 126 | 1311 | 998 |
| 32 | 43 | 176 | 65 | 30 | 63 | 114 | 434 | 94 | 1899 | 90 | 754 | 417 |
| 33 | 48 | 79 | 112 | 32 | 12 | 96 | 361 | 201 | 1812 | 18 | 452 | 610 |
| 34 | 15 | 154 | 159 | 40 | 26 | 61 | 196 | 140 | 928 | 32 | 233 | 555 |
| 35 | 24 | 252 | 74 | 14 | 33 | 63 | 172 | 158 | 341 | 9 | 282 | 118 |
| 36 | 40 | 134 | 134 | 45 | 20 | 2 | 149 | 101 | 418 | 0 | 22 | 284 |
| 37 | 40 | 201 | 127 | 58 | 14 | 50 | 126 | 40 | 253 | 0 | 52 | 16 |
| 38 | 31 | 230 | 27 | 41 | 6 | 33 | 47 | 2 | 61 | 0 | 9 | 0 |
| 39 | 31 | 107 | 19 | 0 | 8 | 12 | 16 | 0 | 126 | 0 | 0 | 0 |
| 40 | 16 | 93 | 29 | 3 | 0 | 0 | 17 | 0 | 0 | 0 | 0 | 0 |
| 41 | 7 | 12 | 12 | 0 | 0 | 12 | 15 | 0 | 0 | 0 | 0 | 0 |
| 42 | 2 | 0 | 0 | 0 | 0 | 16 | 0 | 0 | 0 | 0 | 0 | 0 |
| 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 44 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 45 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sum | 1761 | 3623 | 10515 | 3237 | 1466 | 6956 | 8411 | 22505 | 37898 | 485537325 | 42273 |  |
| $1-$ group | 1388 | 1171 | 4442 | 1239 | 262 | 1108 | 782 | 2058 | 3885 | 873 | 13016 | 22653 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 4.6.1.12.1 Catch in number by age of blue whiting from the summer (August/September) bottom trawl surveys on the Faroe plateau 1996-2004. Icelandic age readings from June-August) were used to split the numbers by age.

|  |  |  |  | Age |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | Total |
| 1996 | 12513 | 18586 | 4576 | 5392 | 6754 | 2755 | 1610 | 768 | 352 | 337 | 121 | 2 | 34 | 53802 |
| 1997 | 4139 | 20745 | 13710 | 8345 | 5748 | 2488 | 1376 | 619 | 242 | 179 | 95 | 2 | 14 | 57701 |
| 1998 | 2359 | 21202 | 28278 | 19217 | 12289 | 4143 | 2330 | 1057 | 358 | 301 | 126 | 4 | 27 | 91690 |
| 1999 | 7322 | 4189 | 4468 | 12725 | 19609 | 6041 | 791 | 524 | 344 | 284 | 139 | 0 | 18 | 56452 |
| 2000 | 11120 | 85876 | 18307 | 18875 | 42059 | 10892 | 2557 | 584 | 270 | 400 | 316 | 0 | 0 | 191254 |
| 2001 | 17431 | 65857 | 49449 | 16099 | 25119 | 9486 | 3362 | 1295 | 420 | 134 | 0 | 0 | 0 | 188652 |
| 2002 | 1113 | 12348 | 10026 | 7112 | 5623 | 5724 | 3616 | 1577 | 448 | 508 | 0 | 0 | 0 | 48095 |
| 2003 | 60646 | 18043 | 17338 | 21706 | 12578 | 4791 | 3701 | 1424 | 357 | 49 | 0 | 9 | 0 | 140641 |
| 2004 | 35744 | 18243 | 10222 | 16912 | 20938 | 6887 | 823 | 550 | 287 | 315 | 137 | 0 | 76 | 111133 |

Table 4.6.1.14.1. Abundance index on blue whiting in the Norwegian winter survey (late January-early March) in the Barents Sea. Blue whiting $\leq 20 \mathrm{~cm}$ in total body length most likely belong to 1 -group.

| Year | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Catch rate | All | 0.10 | 0.24 | 5.41 | 7.66 | 38.3 | 24.2 | 11.8 | 6.57 | 1.88 | 18.4 | 56.1 |
| (ind. nm ) | $<20 \mathrm{~cm}$ | 0.00 | 0.01 | 4.47 | 6.63 | 1.08 | 1.00 | 0.01 | 0.06 | 0.45 | 12.1 | 12.1 |


| Year |  | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Catch rate | All | 3.7 | 1.9 | 23.3 | 220.7 | 20.2 | 8.1 | 106.3 | 396.0 | 119.6 | 70.4 | 171.1 |
| (ind. $/ \mathrm{nm}$ ) | $<20 \mathrm{~cm}$ | 0.00 | 0.17 | 21.07 | 208.05 | 8.76 | 1.71 | 66.54 | 313.79 | 38.07 | 4.46 | 65.02 |

Table 4.6.1.15.1 Blue whiting egg number and development stages in the samples west of British Isles in March-April 2005

| Longitude | Latitude | nos. egg/m ${ }^{2}$ | Egg development stages |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 1 | 2 | 3 | 4 | 5 | 6 |
| 17.51 | 54.62 | 0 |  |  |  |  |  |  |
| 16.00 | 55.38 | 0 |  |  |  |  |  |  |
| 18.00 | 55.37 | 0 |  |  |  |  |  |  |
| 18.34 | 55.87 | 0 |  |  |  |  |  |  |
| 16.50 | 55.88 | 76 | 4 | 3 | 16 | 6 | 20 | 20 |
| 17.09 | 56.13 | 166 | 56 | 20 | 44 | 14 | 20 | 12 |
| 16.00 | 56.37 | 94 | 10 | 2 | 8 | 16 | 20 | 38 |
| 18.68 | 56.37 | 0 |  |  |  |  |  |  |
| 17.92 | 56.63 | 26 | 2 | 6 | 12 | 6 |  |  |
| 17.00 | 57.12 | 94 | 6 | 8 | 34 | 38 | 6 | 2 |
| 17.67 | 57.12 | 2 |  |  |  |  |  |  |
| 15.49 | 57.38 | 14 |  |  |  | 4 | 6 | 4 |
| 16.03 | 57.64 | 40 | 2 | 4 | 2 | 8 | 18 | 6 |
| 17.03 | 57.65 | 2 |  |  |  |  |  |  |
| 15.00 | 58.13 | 4 |  |  |  |  | 4 |  |
| 16.03 | 58.13 | 280 |  | 7 | 7 |  | 95 | 171 |
| 17.03 | 58.13 | 32 | 2 |  |  | 2 | 10 | 18 |
| 17.76 | 58.38 | 2 |  |  |  | 2 |  |  |
| 15.00 | 58.64 | 0 |  |  |  |  |  |  |
| 16.00 | 58.63 | 1582 | 37 | 38 | 217 | 198 | 993 | 99 |
| 17.00 | 58.64 | 34 | 2 | 2 | 4 | 18 | 4 |  |
| 17.85 | 58.87 | 2 |  | 2 |  |  |  |  |
| 14.66 | 59.14 | 224 | 62 | 18 | 85 | 28 | 18 | 13 |
| 16.00 | 59.12 | 160 | 2 | 2 |  | 12 | 44 | 8 |
| 16.50 | 59.50 | 0 |  |  |  |  |  |  |
| 14.14 | 59.63 | 220 | 128 | 18 | 13 | 5 | 4 | 9 |
| 14.94 | 59.62 | 0 |  |  |  |  |  |  |
| 15.97 | 59.62 | 0 |  |  |  |  |  |  |
| 15.33 | 59.87 | 2 |  |  |  |  |  |  |
| 13.48 | 60.12 | 84 |  | 6 | 8 | 24 | 32 | 14 |
| 16.00 | 53.00 | 0 |  |  |  |  |  |  |
| 15.52 | 53.00 | 4 |  |  |  | 2 | 2 |  |
| 15.00 | 53.00 | 2 |  |  |  |  | 2 |  |
| 14.25 | 53.77 | 0 |  |  |  |  |  |  |
| 15.13 | 53.76 | 0 |  |  |  |  |  |  |
| 15.50 | 54.23 | 2 |  |  |  |  | 2 |  |
| 15.50 | 54.75 | 0 |  |  |  |  |  |  |
| 14.49 | 54.74 | 0 |  |  |  |  |  |  |
| 11.75 | 54.75 | 8 |  |  | 8 |  |  |  |
| 11.33 | 55.25 | 2 | 2 |  |  |  |  |  |
| 14.00 | 55.25 | 6 |  |  |  | 2 | 4 |  |
| 14.98 | 55.25 | 0 |  |  |  |  |  |  |
| 15.00 | 55.76 | 32 | 4 |  | 2 | 2 | 14 | 10 |
| 13.83 | 55.75 | 2 |  |  |  |  | 2 |  |
| 12.67 | 55.75 | 0 |  |  |  |  |  |  |
| 11.48 | 55.75 | 0 |  |  |  |  |  |  |
| 11.50 | 56.25 | 74 | 8 | 2 | 2 | 28 | 30 | 4 |
| 12.84 | 56.34 | 4 |  |  | 2 | 2 |  |  |
| 13.83 | 56.33 | 8 |  |  | 2 | 4 | 2 |  |
| 13.51 | 58.24 | 100 |  | 2 | 12 | 32 | 32 | 22 |
|  |  | Total | 327 | 140 | 478 | 453 | 1384 | 450 |

Table 4.7.4.1. Blue whiting. Description of ICA runs performed at the WG in 2005. Mean values of SSB and F are presented for comparison.

| Ages |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Run | parable peri | removed | Jwn weighti | No. fleets | Ref. age | loitation pa | 1 weigh | F | Mean F | SSB | SSB Mean |
|  | 33 | No | Manual | 1 |  | 1 period | 1 | 0.4 | 0.29 | 7.4 | 3.3 |
|  | 43 | No | Manual | 1 | 3 | 1 period | 0.5 | 0.4 | 0.3 | 7.3 | 3.2 |
|  | 54 | No | Manual | 6 | 5 | 1 period | 0.5 | 0.46 | 0.43 | 6.3 | 2.9 |
|  | 18 | No | Manual | 6 | 3 | 1 period | 0.5 | 0.46 | 0.33 | 5.9 | 2.3 |
| $2^{*}$ | * 8 | No | Manual | 1 | 3 | 1 period | 0.5 | 0.46 | 0.33 | 5.8 | 2.9 |
| Final 2004 | 4 | No | Manual | 6 | 3 | 1 period | 0.5 | 0.37 | 0.3 | 6.1 | 2.8 |

* Final 2005

Table 4.7.5.1 Blue Whiting, SMS exploratory run. Estimated coefficient of variation for catch and survey observations using the same setting as for last year's WG, input data for 2004 and additional survey observations for 2005.


## Table 4.7.5.2 Blue Whiting, SMS exploratory run. Diagnostics

```
objective function, negative log likelihood: -146.552
objective function weight:
catch survey Stock/recruit
    1 1 0.01
objective function contributions (total):
                            Catch CPUE S/R
    -148.9 2.3 7.6
objective function contributions (per observation):
    Catch CPUE S/R
    -0.62 0.01 0.31
```

contribution by fleets:

| Spawning grounds 1981-1990 | total: 8.502 | mean: 0.152 |
| :--- | :--- | :--- |
| Spawning grounds 1991-2005 | total:-4.850 | mean:-0.049 |
| Norw. Sea May 2000-2005 | total:-1.371 | mean:-0.125 |

F, Year effect:
1981: 1.000
1982: 0.803
1983: 1.014
1984: 1.275
1985: 1.398
1986: 1.708
1987: 1.360
1988: 1.296
1989: 1.688
1990: 1.655
1991: 0.741
1992: 0.723
1993: 1.000
1994: 0.892
1995: 1.195
1996: 1.649
1997: 1.631
1998: 2.212
1999: 1.847
2000: 2.332
2001: 2.139
2002: 2.019
2003: 2.135
2004: 2.477
F, age effect:

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $1981-1992:$ | 0.074 | 0.114 | 0.169 | 0.217 | 0.265 | 0.356 | 0.414 | 0.430 | 0.430 | 0.430 |
| $1993-2004:$ | 0.036 | 0.055 | 0.135 | 0.200 | 0.214 | 0.244 | 0.242 | 0.269 | 0.269 | 0.269 |

Table 4.7.5.2 (continued) Blue Whiting, SMS exploratory run. Diagnostics

| sqrt(catch variance) $\sim$ CV: |  |
| :--- | :--- |
| ----------------------- |  |
| age |  |
| 1 | 0.377 |
| 2 | 0.339 |
| 3 | 0.250 |
| 4 | 0.250 |
| 5 | 0.250 |
| 6 | 0.250 |
| 7 | 0.493 |
| 8 | 0.493 |
| 9 | 0.493 |
| 10 | 0.493 |


Norw. Sea May 2000-2005 0.520 .55

Table 4.7.5.3 Blue Whiting, SMS run. Estimated fishing mortality using the survey on the spawning grounds and the surveys for juveniles (Norwegian Sea May 2000-2005).

| Age |  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 0.0745 | 0.0598 | 0.0756 | 0.0950 | 0.1041 | 0.1272 | 0.1013 | 0.0965 |
|  | 2 | 0.1140 | 0.0916 | 0.1156 | 0.1454 | 0.1594 | 0.1947 | 0.1551 | 0.1477 |
|  | 3 | 0.1688 | 0.1356 | 0.1712 | 0.2153 | 0.2360 | 0.2884 | 0.2296 | 0.2187 |
|  | 4 | 0.2174 | 0.1747 | 0.2205 | 0.2772 | 0.3039 | 0.3714 | 0.2957 | 0.2817 |
|  | 5 | 0.2646 | 0.2125 | 0.2683 | 0.3373 | 0.3698 | 0.4519 | 0.3598 | 0.3428 |
|  | 6 | 0.3557 | 0.2858 | 0.3608 | 0.4536 | 0.4972 | 0.6076 | 0.4838 | 0.4609 |
|  | 7 | 0.4143 | 0.3328 | 0.4202 | 0.5282 | 0.5790 | 0.7076 | 0.5635 | 0.5367 |
|  | 8 | 0.4299 | 0.3454 | 0.4360 | 0.5481 | 0.6008 | 0.7342 | 0.5847 | 0.5569 |
|  | 9 | 0.4299 | 0.3454 | 0.4360 | 0.5481 | 0.6008 | 0.7342 | 0.5847 | 0.5569 |
|  | 10 | 0.4299 | 0.3454 | 0.4360 | 0.5481 | 0.6008 | 0.7342 | 0.5847 | 0.5569 |
| Avg. F |  | 0.284 | 0.228 | 0.288 | 0.362 | 0.397 | 0.485 | 0.386 | 0.368 |
| Age |  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
|  | 1 | 0.1257 | 0.1233 | 0.0552 | 0.0539 | 0.0359 | 0.0320 | 0.0429 | 0.0592 |
|  | 2 | 0.1924 | 0.1887 | 0.0845 | 0.0824 | 0.0554 | 0.0494 | 0.0662 | 0.0913 |
|  | 3 | 0.2850 | 0.2794 | 0.1251 | 0.1221 | 0.1352 | 0.1205 | 0.1615 | 0.2228 |
|  | 4 | 0.3670 | 0.3598 | 0.1611 | 0.1572 | 0.2004 | 0.1787 | 0.2394 | 0.3303 |
|  | 5 | 0.4465 | 0.4379 | 0.1961 | 0.1913 | 0.2139 | 0.1908 | 0.2555 | 0.3526 |
|  | 6 | 0.6004 | 0.5888 | 0.2637 | 0.2572 | 0.2442 | 0.2178 | 0.2917 | 0.4026 |
|  | 7 | 0.6992 | 0.6857 | 0.3071 | 0.2995 | 0.2424 | 0.2162 | 0.2896 | 0.3997 |
|  | 8 | 0.7255 | 0.7115 | 0.3186 | 0.3108 | 0.2686 | 0.2396 | 0.3209 | 0.4428 |
|  | 9 | 0.7255 | 0.7115 | 0.3186 | 0.3108 | 0.2686 | 0.2396 | 0.3209 | 0.4428 |
|  | 10 | 0.7255 | 0.7115 | 0.3186 | 0.3108 | 0.2686 | 0.2396 | 0.3209 | 0.4428 |
| Avg. F | 3-7 | 0.480 | 0.470 | 0.211 | 0.205 | 0.207 | 0.185 | 0.248 | 0.342 |
| Age |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
|  | 1 | 0.0585 | 0.0794 | 0.0663 | 0.0837 | 0.0768 | 0.0725 | 0.0767 | 0.0889 |
|  | 2 | 0.0903 | 0.1225 | 0.1023 | 0.1291 | 0.1185 | 0.1118 | 0.1183 | 0.1372 |
|  | 3 | 0.2204 | 0.2990 | 0.2496 | 0.3151 | 0.2892 | 0.2729 | 0.2886 | 0.3347 |
|  | 4 | 0.3267 | 0.4432 | 0.3700 | 0.4672 | 0.4287 | 0.4046 | 0.4279 | 0.4962 |
|  | 5 | 0.3487 | 0.4731 | 0.3950 | 0.4987 | 0.4576 | 0.4318 | 0.4567 | 0.5297 |
|  | 6 | 0.3982 | 0.5402 | 0.4510 | 0.5694 | 0.5224 | 0.4930 | 0.5214 | 0.6048 |
|  | 7 | 0.3953 | 0.5363 | 0.4477 | 0.5653 | 0.5187 | 0.4895 | 0.5177 | 0.6004 |
|  | 8 | 0.4380 | 0.5942 | 0.4961 | 0.6263 | 0.5747 | 0.5424 | 0.5736 | 0.6653 |
|  | 9 | 0.4380 | 0.5942 | 0.4961 | 0.6263 | 0.5747 | 0.5424 | 0.5736 | 0.6653 |
|  | 10 | 0.4380 | 0.5942 | 0.4961 | 0.6263 | 0.5747 | 0.5424 | 0.5736 | 0.6653 |
| Avg. F | 3-7 | 0.338 | 0.458 | 0.383 | 0.483 | 0.443 | 0.418 | 0.442 | 0.513 |

Table 4.7.5.4 Blue Whiting, SMS run. Estimated stock numbers and biomass using the survey on the spawning grounds and the surveys for juveniles (Norwegian Sea May 2000-2005).

| Age |  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 3342387 | 3858693 | 14117028 | 18885537 | 11256144 | 8916707 | 9852958 | 7195757 |
|  | 2 | 3629317 | 2540067 | 2975701 | 10716999 | 14061105 | 8304474 | 6428173 | 7289620 |
|  | 3 | 4445215 | 2651242 | 1897600 | 2170261 | 7587147 | 9816400 | 5596020 | 4506933 |
|  | 4 | 2511825 | 3074039 | 1895316 | 1309128 | 1432720 | 4906090 | 6023593 | 3641595 |
|  | 5 | 2277662 | 1654651 | 2113441 | 1244687 | 812322 | 865620 | 2770771 | 3669191 |
|  | 6 | 2251836 | 1431309 | 1095315 | 1323139 | 727282 | 459494 | 451051 | 1582961 |
|  | 7 | 1852758 | 1291769 | 880551 | 625169 | 688274 | 362169 | 204903 | 227635 |
|  | 8 | 1754642 | 1002395 | 758188 | 473614 | 301808 | 315813 | 146133 | 95494 |
|  | 9 | 1503333 | 934636 | 581025 | 401407 | 224146 | 135501 | 124083 | 66677 |
|  | 10 | 3109902 | 2457308 | 1966097 | 1348520 | 828184 | 472459 | 238867 | 165605 |
| TSB |  | 3248753 | 2638905 | 2115739 | 2156721 | 2640955 | 2845648 | 2543199 | 2344468 |
| SSB |  | 2930993 | 2376228 | 1865496 | 1566284 | 1842102 | 2183449 | 1958717 | 1806309 |
| Age |  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
|  | 1 | 9233757 | 23931786 | 8249899 | 5784619 | 5412704 | 6115437 | 8017621 | 23367810 |
|  | 2 | 5349369 | 6666769 | 17320888 | 6391610 | 4487726 | 4275274 | 4849107 | 6288699 |
|  | 3 | 5148656 | 3613021 | 4519629 | 13031968 | 4818957 | 3476296 | 3331601 | 3715964 |
|  | 4 | 2965004 | 3170147 | 2236938 | 3265100 | 9443669 | 3446658 | 2522922 | 2320999 |
|  | 5 | 2249572 | 1681896 | 1811095 | 1558866 | 2284405 | 6328001 | 2360074 | 1625904 |
|  | 6 | 2132331 | 1178474 | 888744 | 1218770 | 1054106 | 1510180 | 4281128 | 1496595 |
|  | 7 | 817438 | 957734 | 535504 | 558996 | 771562 | 676046 | 994440 | 2618248 |
|  | 8 | 108964 | 332605 | 395005 | 322514 | 339215 | 495703 | 445864 | 609450 |
|  | 9 | 44797 | 43185 | 133687 | 235165 | 193519 | 212305 | 319382 | 264840 |
|  | 10 | 108966 | 60941 | 41852 | 104506 | 203814 | 248679 | 297012 | 366134 |
| TSB |  | 2195852 | 1948557 | 3091330 | 3333611 | 3074713 | 2905910 | 2718704 | 2510451 |
| SSB |  | 1744354 | 1510896 | 1977477 | 2634625 | 2533532 | 2438528 | 2257063 | 2065900 |
| Age |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
|  | 1 | 43837166 | 28490055 | 21897867 | 42477405 | 67311890 | 40554444 | 41523647 | 28931287 |
|  | 2 | 18032437 | 33850197 | 21544895 | 16778342 | 31984836 | 51035921 | 30881507 | 31487779 |
|  | 3 | 4699493 | 13489011 | 24518818 | 15924628 | 12072871 | 23261102 | 37364203 | 22463676 |
|  | 4 | 2434693 | 3086643 | 8189913 | 15640206 | 9513661 | 7402448 | 14496293 | 22922145 |
|  | 5 | 1365712 | 1437818 | 1622333 | 4631511 | 8025792 | 5073701 | 4044098 | 7737172 |
|  | 6 | 935624 | 788947 | 733453 | 894835 | 2302930 | 4158211 | 2697235 | 2097064 |
|  | 7 | 819234 | 514429 | 376353 | 382529 | 414578 | 1118241 | 2079316 | 1310985 |
|  | 8 | 1437361 | 451720 | 246350 | 196919 | 177949 | 202062 | 561154 | 1014425 |
|  | 9 | 320446 | 759442 | 204152 | 122816 | 86183 | 82008 | 96179 | 258887 |
|  | 10 | 331763 | 344600 | 498965 | 350534 | 207163 | 135189 | 103384 | 92068 |
| TSB |  | 2968856 | 4878040 | 5369800 | 5228329 | 6249552 | 8399799 | 8978987 | 8144671 |
| SSB |  | 2035801 | 3105755 | 3819928 | 3935254 | 4288840 | 5526287 | 6639399 | 6106538 |

Table 4.7.5.5 Blue Whiting, SMS run. Results using the survey on the spawning grounds and the surveys for juveniles (Norwegian Sea May 2000-2005). For the calculation of SSB in 2005, a recruitment of 43700 millions is assumed.

| Year | $\begin{gathered} \text { Recruits } \\ (1000) \end{gathered}$ | $\begin{gathered} \text { SSB } \\ \text { (tonnes) } \end{gathered}$ | $\begin{gathered} \text { TSB } \\ \text { (tonnes) } \end{gathered}$ | $\begin{gathered} \text { SOP } \\ \text { (tonnes) } \end{gathered}$ | $\begin{array}{r} \text { mean }-F \\ \text { age } 3-7 \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 3342387 | 2930993 | 3248753 | 922980 | 0.284 |
| 1982 | 3858693 | 2376228 | 2638905 | 550643 | 0.228 |
| 1983 | 14117028 | 1865496 | 2115739 | 553344 | 0.288 |
| 1984 | 18885537 | 1566284 | 2156721 | 615569 | 0.362 |
| 1985 | 11256144 | 1842102 | 2640955 | 678214 | 0.397 |
| 1986 | 8916707 | 2183449 | 2845648 | 847145 | 0.485 |
| 1987 | 9852958 | 1958717 | 2543199 | 654718 | 0.386 |
| 1988 | 7195757 | 1806309 | 2344468 | 552264 | 0.368 |
| 1989 | 9233757 | 1744354 | 2195852 | 630316 | 0.480 |
| 1990 | 23931786 | 1510896 | 1948557 | 558128 | 0.470 |
| 1991 | 8249899 | 1977477 | 3091330 | 364008 | 0.211 |
| 1992 | 5784619 | 2634625 | 3333611 | 474592 | 0.205 |
| 1993 | 5412704 | 2533532 | 3074713 | 475198 | 0.207 |
| 1994 | 6115437 | 2438528 | 2905910 | 457696 | 0.185 |
| 1995 | 8017621 | 2257063 | 2718704 | 505176 | 0.248 |
| 1996 | 23367810 | 2065900 | 2510451 | 621104 | 0.342 |
| 1997 | 43837166 | 2035801 | 2968856 | 639681 | 0.338 |
| 1998 | 28490055 | 3105755 | 4878040 | 1131955 | 0.458 |
| 1999 | 21897867 | 3819928 | 5369800 | 1261033 | 0.383 |
| 2000 | 42477405 | 3935254 | 5228329 | 1412449 | 0.483 |
| 2001 | 67311890 | 4288840 | 6249552 | 1771805 | 0.443 |
| 2002 | 40554444 | 5526287 | 8399799 | 1556955 | 0.418 |
| 2003 | 41523647 | 6639399 | 8978987 | 2365319 | 0.442 |
| 2004 | 28931287 | 6106538 | 8144671 | 2383504 | 0.513 |
| 2005 |  | 6107000 |  |  |  |

Table 4.7.7.1 Tuning data for the blue whiting assessment. Inside the framed areas constant selection pattern is assumed. $-1=$ missing data.


Norwegian acoustic spawning stock survey, ages 2-8

|  |  |  | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 3 | 1 | 2372 | 7583 | 3253 | 3647 | 4611 | 4638 | 3654 |
| 1982 | 3 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1983 | 3 | 1 | 297 | 2108 | 2723 | 6511 | 3735 | 3650 | 3153 |
| 1984 | 3 | 1 | 15767 | 1721 | 1616 | 1719 | 1858 | 1128 | 567 |
| 1985 | 3 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1986 | 3 | 1 | 1003 | 5829 | 4122 | 624 | 228 | 203 | 250 |
| 1987 | 3 | 1 | 4960 | 8417 | 22589 | 4735 | 282 | 417 | 385 |
| 1988 | 3 | 1 | 9712 | 9090 | 12367 | 20392 | 7355 | 723 | 599 |
| 1989 | 3 | 1 | 6787 | 22270 | 9973 | 10504 | 7803 | 933 | 293 |
| 1990 | 3 | 1 | 14169 | 12670 | 11228 | 5587 | 6556 | 3273 | 516 |
| 1991 | 3 | 1 | 11147 | 6340 | 8497 | 7407 | 4558 | 2019 | 545 |
| 1992 | 3 | 1 | 1232 | 26123 | 4719 | 1574 | 1386 | 810 | 616 |
| 1993 | 3 | 1 | 4489 | 3321 | 26771 | 2643 | 1270 | 557 | 426 |
| 1994 | 3 | 1 | 1603 | 2950 | 4476 | 11354 | 1742 | 1687 | 908 |
| 1995 | 3 | 1 | 8538 | 9874 | 7906 | 6861 | 9467 | 1795 | 1083 |
| 1996 | 3 | 1 | 8781 | 7433 | 8371 | 2399 | 4455 | 4111 | 1202 |
| 1997 | 3 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1998 | 3 | 1 | 18218 | 34991 | 4697 | 1674 | 279 | 407 | 381 |
| 1999 | 3 | 1 | 19034 | 60309 | 26103 | 1481 | 316 | 72 | 153 |
| 2000 | 3 | 1 | 8613 | 31011 | 41382 | 6843 | 898 | 427 | 228 |
| 2001 | 3 | 1 | 44162 | 12843 | 13805 | 8292 | 718 | 175 | 51 |
| 2002 | 3 | 1 | 71996 | 54740 | 12757 | 5266 | 8404 | 1450 | 305 |
| 2003 | 3 | 1 | 23992 | 70303 | 28756 | 5735 | 2430 | 1708 | 260 |
| 2004 | 3 | 1 | 18569 | 40669 | 50137 | 15649 | 4454 | 2218 | 1313 |
| 2005 | 3 | 1 | 1456 | 19968 | 30459 | 31708 | 7455 | 1993 | 747 |

Norwegian Sea ecosystem survey, ages 1-2

|  |  | 1 | 2 |  |
| :--- | :--- | :--- | ---: | ---: |
| 2000 | 5 | 1 | 182 | 728 |
| 2001 | 5 | 1 | 184 | 460 |
| 2002 | 5 | 1 | 22356 | 396 |
| 2003 | 5 | 1 | 30380 | 13916 |
| 2004 | 5 | 1 | 5969 | 23876 |
| 2005 | 5 | 1 | 2324 | 2380 |

Table 4.9.1.1 Blue Whiting. Input to short term projection.

| Age | Weight in the stock | Weight in the catch | Proportion <br> mature | Natural mortality | F status quo | Stock numbers <br> 2005 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $(\mathrm{~kg})$ | $(\mathrm{kg})$ |  |  |  | (millions) |
| 1 | 0.0488 | 0.0488 | 0.11 | 0.2 | 0.089 | 43700.0 |
| 2 | 0.0732 | 0.0732 | 0.40 | 0.2 | 0.137 | 21671.8 |
| 3 | 0.0935 | 0.0935 | 0.82 | 0.2 | 0.335 | 22476.0 |
| 4 | 0.1083 | 0.1083 | 0.86 | 0.2 | 0.496 | 13159.9 |
| 5 | 0.1288 | 0.1288 | 0.91 | 0.2 | 0.530 | 11425.9 |
| 6 | 0.1530 | 0.1530 | 0.94 | 0.2 | 0.605 | 3729.8 |
| 7 | 0.1718 | 0.1718 | 1.00 | 0.2 | 0.600 | 937.8 |
| 8 | 0.1980 | 0.1980 | 1.00 | 0.2 | 0.665 | 588.8 |
| 9 | 0.2240 | 0.2240 | 1.00 | 0.2 | 0.665 | 427.0 |
| $10+$ | 0.2775 | 0.2775 | 1.00 | 0.2 | 0.665 | 147.7 |

Recruitment in 2006: Geometric mean 1996-2003. 36400 millions, $\mathrm{CV}=37 \%$.

Table 4.9.1.2 Blue Whiting. Short term projection results.
Option a) Catch in 2005 is assumed at 2 million tonnes

Deterministic projection


MCMC projection


Confidence intv. corresponds to the 25 - 75 percentiles
Option b) Catch in 2005 from F status quo
Deterministic projection


MCMC projection



Figure 4.3.1. Total catches of blue whiting in 2004 by ICES rectangle. Grading of the symbols: small dots 10 100 t , white squares 100-1 000 t , grey squares $1000-10000 \mathrm{t}$, and black squares $>10000 \mathrm{t}$.


Figure 4.3.2. Total catches of blue whiting in 2004 by quarter and ICES rectangle. Grading of the symbols: small dots 10-100 t, white squares 100-1 000 t , grey squares $1000-10000 \mathrm{t}$, and black squares $>10000 \mathrm{t}$.


Fig. 4.4.1.1. Geographic distribution of saithe (Pollachius virens) by-catch rates (\% weight) in sampled hauls 2004. (a) May-June, (b) July, (c) August-September, (d) October, (e) November, (f) December (Pálsson et al. 2004).


Fig. 4.4.1.2. Geographic distribution of cod (Gadus morhua) by-catch rates (\% weight) in sampled hauls by months 2004. (a) May-June, (b) July, (c) August-September, (d) October, (e) November, (f) December (Pálsson et al. 2004).


Fig. 4.4.1.3. Length distributions (numbers) of by-catch of (a) cod and (b) saithe in the Icelandic blue whiting fishery 2004 (Pálsson et al. 2004).


Figure 4.4.2.1. Faroese screenings on board the factory trawler M/T Næraberg in November-December 2004 (Lamhauge 2004).


Figure 4.4.2.2. Length distribution (percentage) in cm of by-catch of saithe caught in the screenings on board the factory trawler M/T Næraberg in November-December 2004 (Lamhauge 2004).


Figure 4.5.2.1. Development of blue whiting fisheries in different sub-areas in terms of absolute (top) and relative catches (bottom).

## CPUE Spanish pair trawlers



Figure 4.5.9.1.1. Blue whiting CPUE from Spanish Pair trawlers in ICES Div VIIIc and IXa (North).


Figure 4.6.1.1.1. Blue Whiting: Schematic map of blue whiting acoustic density ( $\mathrm{s}_{\mathrm{A}}, \mathrm{m}^{2} / \mathrm{nm}$ ) in the international spring survey in 2005.


Figure 4.6.1.2.1. Length and age distribution of blue whiting estimated from the Norwegian blue whiting spawning stock survey with R/V "G. O. Sars" in March-April 2005.


Figure 4.6.1.4.1. Density of blue whiting in terms of $\mathrm{s}_{\mathrm{A}}$-values $\left(\mathrm{m}^{2} / \mathrm{nm}^{2}\right)$ based on combined 5 nm values reported by each of the research vessels "Dana", "Magnus Heinason", "Arni Fridriksson", "Johan Hjort" and "G. O. Sars" in the Norwegian Sea-Faroese EEZ in May 2005.


Fig. 4.6.1.8.1 Blue whiting distribution in October-December $2004\left(\mathrm{~s}_{\mathrm{A}}, \mathrm{m}^{2} / \mathrm{mile}^{2}\right)$.

## Spanish Bottom Trawl Survey



Spanish Bottom Trawl Survey


Figure 4.6.1.9.1 Blue whiting. Mean catch rates (Kg/haul and Number/haul) in Spanish bottom trawl survey.

## Portuguese bottom trawl survey (Summer)



## Portuguese bottom trawl survey (Autumn)



Figure 4.6.1.10.1 Blue whiting. Mean catch rates in the Portuguese bottom trawl surveys.


Figure 4.6.1.12.1 Bottom trawl 0 -group index (number of blue whiting caught $10^{-3}$ ) of Faroes plateau bottom trawl survey from autumn 1996-2004 compared to the following 1-group spring index lagged one year to match the 0 -group index. Year is indicated on the points.


Fig. 4.6.1.15.1. Number of blue whiting eggs above $1 \mathrm{~m}^{\wedge} 2$ of the area investigated by R/V ATLANTNIRO 16.03-08.04.2005 and R/V FRIDTJOF NANSEN 18-29.03.2005


Figure 4.7.1.1 Age disaggregated catch in numbers of blue whiting plotted on log scale. The labels above each figure indicate yearclasses. The grey lines correspond to $\mathrm{Z}=0.6$.


Figure 4.7.1.2. Abundance index of blue whiting in the spawning fish survey plotted against the abundance index of the same yearclass the following year. The dashed line shows the most recent estimate. The labels in the picture indicate yearclass.


Figure 4.7.1.3 Age disaggregated abundance indices of blue whiting from the spawning survey plotted on log scale. The labels above each figure indicate yearclasses. The grey lines correspond to $\mathrm{Z}=0.6$.


Figure 4.7.1.4. Abundance index of blue whiting in Norwegian sea survey plotted against the abundance index of the same yearclass the following year. The wide line shows the most recent estimate. The labels in the picture indicate yearclass


Figure 4.7.1.5 Number of age 1 blue whiting in the Barents sea survey plotted against estimate from assessment. The lines show curve fitted to the data and the horizontal lines the most recent points.


Figure 4.7.1.6 Some recruitment indices of blue whiting plotted against each other. In the upper 2 figures the label indicates year but yearclass in the lower 2.


Figure 4.7.1.7 Indices of age 1 blue whiting from the Icelandic groundfish survey in March. The indices are calculated as the mean number of fish smaller than 21 cm in each tow using the area covered by the survey in all years.


Figure 4.7.1.8 Blue whiting. Prediction of recruitment from a model $N_{1}=\alpha I_{i c e}+\beta I_{b a r}$. Labels in the figure indicate year. The horizontal line shows the estimate for the 2005 yearclass and the slope line 1-1 relationship so points on that line fit perfectly.


Figure 4.7.2.1. Blue whiting. Comparison of the final AMCI assessment in 2004 and new assessment with the same settings but updated data (SPALY: the same procedure as the last year).


Figure 4.7.2.2. Blue whiting. Residuals from the final AMCI run. In the lower panel, residuals from two surveys are plotted in the same figure: Norwegian spawning stock survey (ages 3-8 years) and international May survey (ages 1-2 years).




Figure 4.7.2.3. Blue whiting. Retrospective analysis on the final AMCI assessment. The terminal year is varied from 2000 to 2005.


Figure 4.7.2.4. Blue whiting. Results from bootstrapping the final AMCI assessment (1000 replicates).


Figure 4.7.2.5. Blue whiting. Results from bootstrapping the final AMCI assessment ( 1000 replicates). These graphs show interdependence between the estimate of fishing mortality in the final year and SSB in the end of the final year. SSB in the end of the final year is also strongly correlated with recruitment in 2001.










## 0 - fit on catch-at-age

1 - fit on Norway spawning acoustic N(a,y) (1981-1990)
2 - fit on Norway spawning acoustic N(a,y( (1991-)
3 - fit on Russian spawning acoustic $\mathrm{N}(\mathrm{a}, \mathrm{y})(1982-1991)$
4 - fit on Russian spawning acoustic $N(a, y)(1992-)$
5 - fit on Norwegian sea acoustic $N(a, y)(1981-1990)$
6 - fit on Norwegian sea acoustic N(a,y) (1991- )
effort-controlled ISVPA (SPALY)

Figure 4.7.3.1 Blue whiting. ISVPA effort-controlled version.










## 0 - fit on catch-at-age

1 - fit on Norway spawning acoustic $\mathrm{N}(\mathrm{a}, \mathrm{y})(1981-1990)$
2 - fit on Norway spawning acoustic $N(a, y($ (1991- )
3 - fit on Russian spawning acoustic $\mathrm{N}(\mathrm{a}, \mathrm{y})$ (1982-1991)
4 - fit on Russian spawning acoustic N(a,y) (1992- ) 5 - fit on Norwegian sea acoustic $N(a, y)(1981-1990)$ 6 - fit on Norwegian sea acoustic $N(a, y)$ (1991- )
catch-controlled ISVPA


Figure 4.7.3.2 Blue whiting. ISVPA, catch-controlled version, 2 sources of information.



Figure 4.7.3.3 Blue whiting. Comparison of ISVPA-derived estimates in 2004 and 2005 assessments.


Figure 4.7.3.4 Blue whiting. Results of ISVPA stock assessment (effort-controlled version, 2 sources of information (catch-at-age and second part of Norwegian spawning acoustic surveys)







Figure 4.7.3.5 Blue whiting, ISVPA, retrospective runs


Figure 4.7.3.6 Blue whiting. Comparison of reported and theoretical (ISVPA-derived) catches in weight for age groups 1-10+.




Figure 4.7.3.7 Blue whiting, ISVPA, bootstrap



Figure 4.7.3.8 Blue whiting. Residuals in catch-at-age and surveys (ISVPA derived)


Figure 4.7.4.1 Blue whiting. Comparison of 5 runs of ICA by WG in 2005. Run 2 is the final run.


Separable Model Diagnostics ue


Prest pumapyint screen, or any ot

| Landings | Fishing Martalitu |
| :---: | :---: |
| Riecruitment | Stack size |

Figure 4.7.4.2. SSq surface plot, catch residuals and selection pattern, and standard plots for final ICA run for blue whiting in 2004.




Figure 4.7.4.3. Residual plots for catches and survey for blue whiting, final run 2 WG 2005.


Figure 4.7.4.4. Blue whiting. Retrospective ICA runs to 2000 using settings for final run 2005



Figure 4.7.5.1 Blue Whiting, SMS exploratory run. Model comparison, SSB and fishing mortality. SSB for 2005 does not include contributions from age 1.


Figure 4.7.5.2 normalized to 1)

## Blue whiting



Figure 4.7.5.3 Blue Whiting, SMS exploratory run, Residuals for catch observations. Red (dark) bubbles show that the observed value is bigger than the expected value.

Spawning grounds 1981-1990


Spawning grounds 1991-2005


Norw. Sea May 2000-2005


Figure 4.7.5.4 Blue Whiting, SMS exploratory run, Residuals for survey observations. Red (dark) bubbles show that the observed value is bigger than the expected value


Figure 4.7.5.5 Blue Whiting, SMS exploratory run, Observed and predicted Yield

## Retrospective 1998-2004



Retrospective 1998-2004


Figure 4.7.5.6 Blue Whiting, SMS exploratory run, Retrospective pattern of F and SSB

## a priori weight on surveys



Figure 4.7.5.7 Blue Whiting, SMS exploratory run. Effect on estimated SSB of varying a priori weight on survey observations. A priori weight on catch observations are kept constant at 1.0. SSB in 2005 does not include contributions from the age 1 .


Figure 4.7.5.8 Blue Whiting, SMS. Stock summary, 1981-2005. For the calculation of SSB in 2005 an recruitment at $43.7^{*} 10^{\wedge} 9$ is assumed.


Figure 4.7.5.9
Blue Whiting, SMS. SSB-recruit plot


Figure 4.7.5.10 Blue Whiting, SMS. Estimated uncertainties on stock numbers.


Figure 4.7.5.11 Blue Whiting, SMS. Estimated uncertainties on F and SSB.

## Blue whiting



Blue whiting


Figure 4.7.5.12 Blue Whiting, SMS. Posterior density (2.5, 25, 50, 75 and 97.5 percentiles) of average $F$ and SSB estimated from 200000 Markov Chain Monte Carlo simulations. Recruitment in 2005 is assumed as the GM of the full time series.

SSB, 2004


F, 2004

$\bar{F}$
95\% confidence interval: 0.389-0.826
density plot SSB F in 2004


SSB, 2005


SSB (1000t)
95\% confidence interval: 3260-7003

Figure 4.7.5.13 Blue Whiting, SMS. Posterior density of SSB, average F and their simultaneous distribution in the 2004 estimated from 200000 Markov Chain Monte Carlo simulations.





Figure 4.7.6.1. Blue Whiting: Comparisons between final AMCI, ISVPA ICA and SMS assessments.


Figure 4.7.6.2 Blue whiting Comparison of $F$ estimated by different tuning methods in WG 2004 and WG 2005



Figure 4.8.1. Overview of survey indices for young blue whiting (Icelandic July acoustic, Barents Sea winter bottom trawl, PGNAPES-coordinated Norwegian Sea May, Faroese bottom trawl spring and summer, and Icelandic winter bottom trawl surveys). Each time series is normalized to zero mean and unity variance. The dotted line shows arithmetic mean of the data - please note that ignores the unequal areal coverage of the surveys.


Figure 4.8.2. Abundance index of blue whiting in Norwegian Sea ecosystem survey in May plotted against the abundance index of the same year-class the following year. The wide line shows the most recent estimate (i.e., 2005). The labels show year-class.



Figure 4.8.3. Number of age 1 year blue whiting in the Barents Sea survey plotted against estimate from the WG assessment in 2004. The lines show curve fitted to the data and the horizontal lines the most recent points (2004-2005).


Figure 4.8.4. Some recruitment indices of blue whiting plotted against each other. The label shows the year of observation.


Figure 4.8.5. Blue whiting. Prediction of recruitment from a model $N_{1}=\alpha I_{i c e}+\beta I_{b a r}$ (i.e., forced through the origin). Labels show the year of observation. The horizontal line shows the estimate for the 2005 yearclass and the diagonal 1:1 relationship (points on that line fit perfectly).


Figure 4.8.6. Blue whiting. Prediction of recruitment from the Barents Sea data calibrated with the assessment estimate from 2004 (upper panel) and 2005 (lower panel). The thin curves show fitted values based on either the whole period and taking into account the change in mesh size in 1994, or on data from 1994 onwards only. Grey horizontal lines show geometric mean WG recruitment for the whole period and for the "high productivity" period from 1996 onwards.

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## Annex 1 - List of participants

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ICES, Headquarters, 25 August - 1 September 2005

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# NPBW Review Group 

28 - 29 September, 2005

Jan Horbowy<br>Asta Gudmundsdottir<br>Mika Kurkilahti<br>John Simmonds

Poland (Chair)<br>Iceland (Chair of WGNPBW)<br>Finland<br>UK Scotland

Technical Minutes

## Norwegian Spring Spawning Herring

Catch at age numbers and weights used in the assessment and presented in the report were different from official catch, this was detected by the WG after the meeting. These were corrected in an ad hoc way to remove SOP differences. None of the SALLOC files for NSSH were on the report directory so it was not possible to trace the cause, the files were obtained late in the meeting and two problems identified, the corrections applied are unlikely to be significant..

The fraction mature in the assessment tables for 2005 differ from those in the projections. During the review group it was confirmed that the fraction mature in the projections was the correct version.

Historic maturity at age looks that it may be incorrect data on 1963-1966 yearclasses at some of the ages 3-6 (very low maturity of 1963 year-class at age 5, in next year-classes maturity at some older ages are lower than at younger ages). Can these be checked and either an explanation or a correction provided

## Assessment.

The WG has run the assessment with settings similar to those from last year which is correct (the update assessment was planed this year) and the assessment is accepted.

A separate assessment and projection was presented by the Russian delegation. It was based on an ISVPA model fit which has poor residual patterns (large year-class effects) and although good retrospective pattern with last year assessment, it had poorer performance in earlier assessment years. In addition, it was noted that in the loss function of ISVPA functionally different terms were used for some of the surveys and statistical effects of such usage are not clear. The ISVPA assessment was not supported by the WG and is not accepted by the review group.

There are number of issues regarding the use of data in the assessment. These issues are not sufficient to make the assessment unacceptable but it would be helpful if they could be resolved. It is therefore recommended that these points should be considered as part of a benchmark assessment in 2006. The assessments should be run as far as possible (if the model can converge) with each survey independently thus illustrating the influence of each source of data on the main assessment results and diagnostic. Survey time series that have ceased more than 5 years before terminal year should be
investigated for utility. In particular survey 2, the January over-wintering survey, should be considered for exclusion as it has no longer been carried out since 1999.
Survey 6, the Barents Sea in autumn survey for ages 1 and 2, has very strong influence on the terminal F in ISVPA. The use of a survey on 1 and 2 age herring too strongly influenced F and the abundance for 4 years and older is particularly of concern. While these are useful to make estimates of incoming year-classes they perhaps should not used to dominate the assessment of older ages. This influence of individual surveys on the Seastar results is not specifically documented and more information would be helpful. The choice of whether a survey should be included and the a priori weight it is given should be based on the precision of the assessment and the restrospective bias resulting from its use.

## Catch prediction

The WG presented three options for catch prediction, differing with assumptions on exploitation pattern of strong 2002 year class in 2006. The Review Group supported option 1 of the WG, i.e. assumption that exploitation of 2002 year class in 2006 (4 years old) will be the same as exploitation of 5 years old in that year. The basis for the choice of F on 4 year olds in 2006 is that the 2002 year-class is expected to be distributed spatially along with the older ages and therefore subject to higher F than previous year-classes at age 4. To obtain a TAC that represents the F correctly for older ages (i.e. one that matches the management plan) the selection on this age needs to be increased, the increase is uncertain but the use of F at age 5 is the best estimate available. This option is recommended for use in the projections for ACFM. In addition, it was unclear why $60 \%$ maturity in 2006 was assumed for strong 2002 year class in prediction, this was not well documented in the report. As the year-class is strong this assumption may matter for estimate of SSB in prediction.

## Report information

The historical estimates of fishing mortality at age were missing. Residual plots of the fit for Seastar fitted cohorts should be provided for comparison with plots from ISVPA. Major cohort plots similar to those from Seastar should be provided from ISVPA. A table of the main parameters for ISVPA runs (SSB, Juveniles, terminal Fs, catchabilities of surveys, like table 3.7.2.1.1) should be provided for comparison with Seastar. Many of the graphs are provided without axes labels (particularly from Seastar graphs)

## Blue Whiting

The documentation standards for the blue whiting assessments are much better than those for NSSH. The exploratory analysis was very helpful. However, the discussion of the 4 different assessment models fits depended more on the individual author than a concerted review. Thus some aspects were ignored in some discussion but the same issues were raised as major concerns with a different model. The major issue is that the description of the fishery is one with substantial changes in fishing pattern over the last 15 years. Despite this the models were all set with long separable model periods. Only AMCI allowed for inclusion of some limited variations in exploitation patterns from year to year but these could be to low when compared with probable variations resulting from diverse fisheries on the stock.

The recently enhanced international spawning area survey used in the assessment excluded the more recent data, despite indications that this was consistent. The WG or (PGNAPES) should now combine this data into a single series.

There are two areas of concern.
The major adult tuning series (spawning area survey) shows a distinctly different selection pattern between the early (EK400) period and the later (EK500) period. The earlier section shows a flatter selection at old age compared with the later period which shows a distinct decline at age. This difference is not discussed or explained, and it may result in incorrect elevation of numbers at older ages.

The spawning survey time-series expressed as a biomass index shows around a factor of two rise in SSB from the early 90s to the last four years. When fitted most strongly in the SMS model the rise in biomass in nearer a factor of four. The final assessment shows a factor of three. This suggests a distinct conflict between catch data and survey data. The separable constraint that all the models use to try to smooth the data may be resulting in a false impression. The WG should consider a less constrained model, representing the observed changes in the fishery better, possibly by allowing more change in recent years in AMCI, or evaluating the use of XSA, with some flexibility in the selection in recent years.

The WG accepted the assessment with SMS model as a final assessment opposite to previous years assessments which were based on AMCI. No compelling argument was provided by the WG for changing the model and given the above reasons the review group did not support the change in assessment method to the newer more constrained model and decided to recommend the previous method as the assessment. Given concerns about the modelling, the stochastic element was ignored as it was thought to under-represent the uncertainty in the assessment. A deterministic short term projection was run and is provided.

The results of corrected herring assessment and projections and results of blue whiting assessment and projections with AMCI are provided below.

## Norwegian spring spawning herring.

Rescaled catch in numbers for 2004.

| Age | Numbers | Mean weight |
| :---: | :---: | :---: |
| 0 | 0.000 | 0.022 |
| 1 | 0.002 | 0.065 |
| 2 | 0.037 | 0.141 |
| 3 | 0.024 | 0.181 |
| 4 | 0.108 | 0.211 |
| 5 | 0.591 | 0.240 |
| 6 | 0.818 | 0.271 |
| 7 | 0.135 | 0.303 |
| 8 | 0.129 | 0.347 |
| 9 | 0.027 | 0.359 |
| 10 | 0.052 | 0.368 |
| 11 | 0.151 | 0.387 |
| 12 | 0.350 | 0.392 |
| 13 | 0.180 | 0.400 |
| 14 | 0.023 | 0.423 |
| $15+$ | 0.010 | 0.438 |

Table. 3.7.2.1.1 NSSH. Summary of SeaStar exploratory runs.

|  | Default | EstimateM | IncreasedMPlusgroup | TimeTrendLarvae | NoLarvae | Log | 2002InBarents | NoTags | CompatibleISVPA | CorrectingAgeBias |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spawning stock | 6.102 | 5.141 | 6.133 | 5.678 | 5.411 | 6.120 | 6.120 | 6.329 | 4.564 | 6.102 |
| Juvenile stock | 5.006 | 4.364 | 4.983 | 4.701 | 4.633 | 5.111 | 5.111 | 5.125 | 4.318 | 5.006 |
| terminalf83 | 0.198 | 0.217 | 0.198 | 0.209 | 0.208 | 0.220 |  | 0.167 | 0.174 | 0.198 |
| terminalF90 | 0.280 | 0.381 | 0.273 | 0.300 | 0.312 | 0.530 |  | 0.285 | 0.392 | 0.280 |
| terminalF91 | 0.268 | 0.334 | 0.264 | 0.288 | 0.295 | 0.285 |  | 0.257 | 0.367 | 0.268 |
| terminalf92 | 0.278 | 0.353 | 0.273 | 0.300 | 0.308 | 0.274 |  | 0.290 | 0.417 | 0.278 |
| terminalf93 | 0.377 | 0.439 | 0.374 | 0.398 | 0.404 | 0.448 |  | 0.307 | 0.437 | 0.377 |
| terminalf96 | 0.226 | 0.279 | 0.223 | 0.238 | 0.246 | 0.242 |  | 0.217 | 0.292 | 0.226 |
| terminalf97 | 0.187 | 0.223 | 0.185 | 0.195 | 0.200 | 0.187 |  | 0.181 | 0.237 | 0.187 |
| terminalF98 | 0.095 | 0.115 | 0.094 | 0.103 | 0.108 | 0.095 |  | 0.093 | 0.127 | 0.095 |
| terminalf99 | 0.072 | 0.087 | 0.070 | 0.077 | 0.081 | 0.070 |  | 0.070 | 0.090 | 0.072 |
| terminalF02 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 | 0.001 |  | 0.001 | 0.001 | 0.001 |
| terminalF03 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| terminalF04 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| cat1 | 1.053 | 1.257 | 1.057 | 1.104 | 1.113 | 1.036 |  | 1.009 |  | 1.053 |
| cat2 | 0.846 | 1.012 | 0.831 | 0.861 | 0.885 | 0.755 |  | 0.828 | 0.951 | 0.846 |
| cat3 | 0.943 | 1.119 | 0.944 | 0.957 | 0.960 | 0.933 |  | 0.909 |  | 0.943 |
| cat5 | 1.143 | 1.379 | 1.133 | 1.187 | 1.208 | 1.072 |  | 1.102 | 1.326 | 1.143 |
| catLarvae2 | 4.699 | 5.349 | 4.841 | 5.173 |  | 4.748 |  | 4.603 |  | 4.699 |
| cat4 | 0.345 | 0.367 | 0.340 | 0.374 | 0.364 | 0.348 | 0.348 | 0.338 |  | 0.345 |
| cat6 | 0.494 | 0.571 | 0.485 | 0.543 | 0.540 | 0.487 | 0.487 | 0.481 |  | 0.494 |
| catZero | 0.003 | 0.002 | 0.003 | 0.003 | 0.003 | 0.003 | 0.003 | 0.003 | 0.003 | 0.003 |
| distributionParameter1 | 0.486 | 0.478 | 0.491 | 0.494 | 0.485 | 0.525 |  | 0.482 | 0.514 | 0.486 |
| distributionParameterLarvae | e 0.674 | 0.764 | 0.627 | 0.430 |  | 0.674 |  | 0.670 |  | 0.674 |
| distributionParameter2 | 1.945 | 1.822 | 1.210 | 1.227 | 1.948 | 1.205 | 1.205 | 1.948 | 113.989 | 1.945 |
| distParZero | 0.249 | 0.250 | 0.251 | 0.243 | 0.238 | 0.259 | 0.259 | 0.250 | 0.230 | 0.249 |
| Log-likelihood per term, survey 1 | -1.506 | -1.494 | -1.505 | -1.536 | -1.541 | -1.560 |  | -1.475 |  | -1.506 |
| Log-likelihood per term, survey 2 | -1.238 | -1.234 | -1.237 | -1.224 | -1.228 | -1.246 |  | -1.260 | -1.191 | -1.238 |
| Log-likelihood per term, survey 3 | -1.823 | -1.813 | -1.827 | -1.832 | -1.822 | -1.853 |  | -1.835 | -1.191 | -1.823 |
| Log-likelihood per term, survey 4 | -3.453 | -3.432 | -3.280 | -3.277 | -3.452 | -3.276 | -3.276 | -3.457 |  | -3.453 |
| Log-likelihood per term, survey 5 | -1.523 | -1.487 | -1.526 | -1.520 | -1.513 | -1.546 |  | -1.499 | -1.507 | -1.523 |
| Log-likelihood per term, survey 6 | -3.797 | -3.766 | -3.643 | -3.652 | -3.804 | -3.629 | -3.629 | -3.799 |  | -3.797 |
| Number of data points, survey 1 | 27 | 27 | 27 | 27 | 27 | 27 | 0 | 27 | 0 | 27 |
| Number of data points, |  |  |  |  |  |  |  |  |  |  |
| survey 2 | 58 | 58 | 58 | 58 | 58 | 58 | 0 | 58 | 58 | 58 |


| survey 3 |  |  |  |
| :--- | ---: | ---: | ---: |
| Number of data points, <br> survey 4 | 18 | 18 | 18 |
| Number of data points, <br> survey 5 | 20 | 20 | 20 |
| Number of data points, <br> survey 6 | 65 | 65 | 65 |
| vpaM |  |  |  |
| vpaMYoung | 11 | 11 | 11 |
| survLogLik1 | 0.150 | 0.104 | 0.150 |
| survLogLik2 | 0.900 | 1.077 | 0.900 |
| tagLogLik1 | -244.240 | -241.169 | -244.424 |
| tagLogLik2 | -110.823 | -110.061 | -105.663 |
| larvLogLik1 | -369.756 | -365.624 | -369.722 |
| larvLogLik2 | 0.000 | 0.000 | 0.000 |
| zeroLogLik1 | -80.560 | -84.084 | -78.634 |
| zeroLogLik2 | 0.000 | 0.000 | 0.000 |
| predLogLik1 | 0.000 | 0.000 | 0.000 |
| predLogLik2 | -0.362 | -0.104 | -0.460 |
|  | 0.000 | 0.000 | 0.000 |


| 18 | 18 | 18 | 0 |
| ---: | ---: | ---: | ---: |
| 20 | 20 | 20 | 40 |
|  |  |  |  |
| 65 | 65 | 65 | 0 |
|  |  |  |  |
| 11 | 11 | 11 | 22 |
| 0.150 | 0.150 | 0.150 | 0.150 |
| 0.900 | 0.900 | 0.900 | 0.900 |
| -244.198 | -243.987 | -248.199 | -105.438 |
| -105.716 | -110.879 | -105.438 | -105.438 |
| -369.545 | -369.620 | -369.093 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 |
| -68.730 | 0.000 | -80.624 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | -0.870 |
| -0.075 | 0.215 | -0.870 | -0.870 |
| 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 |


| 18 | 0 | 18 |
| ---: | :---: | ---: |
| 20 | 0 | 20 |
|  |  |  |
| 65 | 65 | 65 |
|  |  | 11 |
| 0.150 | 0.150 | 0.150 |
| 0.900 | 0.900 | 0.900 |
| -243.357 | -167.040 | -244.240 |
| -110.919 | 0.000 | -110.823 |
| 0.000 | 0.000 | -369.756 |
| 0.000 | 0.000 | 0.000 |
| -80.387 | 0.000 | -80.560 |
| 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 |
| -0.432 | 0.678 | -0.362 |
| 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 |

Table 3.7.4.1. NSSH. Stock in numbers.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0.000 | 163.151 | 26.549 | 19.391 | 0.921 | 2.070 | 7.492 | 7.709 |
| 2004 | 401.287 | 65.302 | 47.753 | 1.097 | 2.521 | 9.341 | 9.839 | 0.859 |
| 2003 | 160.617 | 117.458 | 2.705 | 3.006 | 11.189 | 12.203 | 1.185 | 0.981 |
| 2002 | 288.899 | 6.652 | 7.492 | 13.214 | 14.871 | 1.652 | 1.492 | 0.223 |
| 2001 | 16.362 | 18.427 | 32.504 | 17.387 | 2.093 | 2.193 | 0.301 | 0.642 |
| 2000 | 45.323 | 79.946 | 42.789 | 2.522 | 3.152 | 0.387 | 0.865 | 2.495 |
| 1999 | 196.635 | 105.244 | 6.211 | 3.810 | 0.488 | 1.150 | 3.362 | 10.105 |
| 1998 | 258.857 | 15.276 | 9.502 | 0.643 | 1.597 | 4.303 | 13.637 | 8.402 |
| 1997 | 37.573 | 23.372 | 1.616 | 1.996 | 5.291 | 17.780 | 11.910 | 2.763 |
| 1996 | 57.485 | 3.974 | 4.958 | 6.184 | 21.427 | 15.531 | 4.224 | 1.519 |
| 1995 | 9.773 | 12.194 | 15.213 | 24.956 | 18.418 | 5.579 | 2.453 | 0.665 |
| 1994 | 29.993 | 37.417 | 61.395 | 21.434 | 6.601 | 3.242 | 0.951 | 0.295 |
| 1993 | 92.042 | 151.007 | 52.729 | 7.700 | 3.882 | 1.199 | 0.352 | 0.212 |
| 1992 | 371.421 | 129.694 | 18.941 | 4.524 | 1.428 | 0.414 | 0.248 | 1.216 |
| 1991 | 318.995 | 46.592 | 11.132 | 1.669 | 0.484 | 0.289 | 1.428 | 0.377 |
| 1990 | 114.598 | 27.381 | 4.129 | 0.583 | 0.339 | 1.672 | 0.449 | 10.405 |
| 1989 | 67.357 | 10.158 | 1.473 | 0.397 | 1.947 | 0.528 | 12.439 | 0.035 |
| 1988 | 25.009 | 3.628 | 0.991 | 2.329 | 0.641 | 15.045 | 0.051 | 0.011 |
| 1987 | 8.945 | 2.447 | 5.786 | 0.766 | 18.020 | 0.079 | 0.017 | 0.021 |
| 1986 | 6.041 | 14.232 | 1.888 | 21.519 | 0.111 | 0.035 | 0.041 | 0.262 |
| 1985 | 35.051 | 4.665 | 53.252 | 0.152 | 0.058 | 0.065 | 0.445 | 0.194 |
| 1984 | 11.528 | 130.982 | 0.378 | 0.072 | 0.081 | 0.583 | 0.246 | 0.168 |
| 1983 | 322.362 | 0.938 | 0.179 | 0.098 | 0.700 | 0.296 | 0.202 | 0.348 |
| 1982 | 2.343 | 0.442 | 0.241 | 0.829 | 0.352 | 0.240 | 0.412 | 0.099 |
| 1981 | 1.100 | 0.595 | 2.057 | 0.413 | 0.283 | 0.487 | 0.118 | 0.279 |
| 1980 | 1.474 | 5.060 | 1.017 | 0.336 | 0.573 | 0.139 | 0.333 | 0.415 |
| 1979 | 12.498 | 2.508 | 0.830 | 0.672 | 0.164 | 0.394 | 0.494 | 0.007 |
| 1978 | 6.201 | 2.044 | 1.655 | 0.194 | 0.471 | 0.596 | 0.009 | 0.000 |
| 1977 | 5.095 | 4.080 | 0.482 | 0.571 | 0.718 | 0.010 | 0.000 | 0.002 |
| 1976 | 10.068 | 1.188 | 1.407 | 0.860 | 0.018 | 0.000 | 0.002 | 0.137 |
| 1975 | 2.971 | 3.467 | 2.117 | 0.024 | 0.000 | 0.004 | 0.192 | 0.000 |
| 1974 | 8.631 | 5.220 | 0.066 | 0.000 | 0.004 | 0.250 | 0.000 | 0.000 |
| 1973 | 12.884 | 0.168 | 0.005 | 0.008 | 0.317 | 0.001 | 0.002 | 0.000 |
| 1972 | 0.957 | 0.077 | 0.051 | 0.407 | 0.005 | 0.006 | 0.003 | 0.001 |
| 1971 | 0.236 | 0.193 | 1.134 | 0.008 | 0.008 | 0.005 | 0.002 | 0.003 |
| 1970 | 0.661 | 3.620 | 0.072 | 0.017 | 0.026 | 0.003 | 0.007 | 0.005 |
| 1969 | 9.785 | 0.972 | 0.263 | 0.233 | 0.004 | 0.018 | 0.011 | 0.003 |
| 1968 | 5.187 | 1.333 | 1.182 | 0.111 | 2.048 | 1.509 | 0.018 | 0.144 |
| 1967 | 3.947 | 18.398 | 0.384 | 3.880 | 5.260 | 0.050 | 0.622 | 1.496 |
| 1966 | 51.409 | 1.984 | 12.175 | 8.320 | 0.087 | 1.225 | 3.146 | 5.897 |
| 1965 | 8.491 | 35.874 | 24.938 | 0.198 | 1.700 | 4.271 | 9.222 | 0.086 |
| 1964 | 93.903 | 65.617 | 0.833 | 2.098 | 5.392 | 12.920 | 0.115 | 0.054 |
| 1963 | 168.931 | 5.371 | 8.368 | 7.085 | 15.912 | 0.139 | 0.064 | 0.058 |


| 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.614 | 0.476 | 0.078 | 0.124 | 0.306 | 1.027 | 0.551 | 0.068 | 0.020 |
| 0.692 | 0.120 | 0.200 | 0.518 | 1.571 | 0.835 | 0.104 | 0.015 | 0.021 |
| 0.164 | 0.312 | 0.831 | 2.418 | 1.199 | 0.166 | 0.027 | 0.009 | 0.029 |
| 0.463 | 1.250 | 3.524 | 1.758 | 0.250 | 0.045 | 0.018 | 0.002 | 0.050 |
| 1.772 | 4.999 | 2.589 | 0.370 | 0.077 | 0.025 | 0.005 | 0.000 | 0.094 |
| 7.208 | 4.135 | 0.664 | 0.167 | 0.046 | 0.031 | 0.025 | 0.026 | 0.158 |
| 6.059 | 1.085 | 0.308 | 0.069 | 0.079 | 0.037 | 0.125 | 0.000 | 0.347 |
| 1.672 | 0.498 | 0.126 | 0.119 | 0.047 | 0.267 | 0.006 | 0.650 | 0.005 |
| 0.931 | 0.213 | 0.160 | 0.089 | 0.408 | 0.028 | 1.155 | 0.001 | 0.006 |
| 0.358 | 0.192 | 0.112 | 0.545 | 0.051 | 2.243 | 0.002 | 0.000 | 0.011 |
| 0.239 | 0.147 | 0.709 | 0.150 | 3.589 | 0.006 | 0.000 | 0.000 | 0.017 |
| 0.179 | 0.863 | 0.212 | 4.866 | 0.010 | 0.000 | 0.000 | 0.023 | 0.006 |
| 1.035 | 0.267 | 6.095 | 0.012 | 0.000 | 0.000 | 0.027 | 0.000 | 0.010 |
| 0.316 | 7.325 | 0.017 | 0.002 | 0.000 | 0.033 | 0.000 | 0.000 | 0.016 |
| 8.746 | 0.022 | 0.003 | 0.000 | 0.039 | 0.000 | 0.000 | 0.024 | 0.002 |
| 0.027 | 0.005 | 0.003 | 0.048 | 0.000 | 0.000 | 0.029 | 0.001 | 0.002 |
| 0.006 | 0.004 | 0.059 | 0.002 | 0.001 | 0.034 | 0.002 | 0.000 | 0.003 |
| 0.011 | 0.085 | 0.012 | 0.005 | 0.043 | 0.005 | 0.002 | 0.002 | 0.003 |
| 0.128 | 0.027 | 0.016 | 0.055 | 0.014 | 0.009 | 0.010 | 0.001 | 0.004 |
| 0.113 | 0.063 | 0.147 | 0.037 | 0.082 | 0.097 | 0.002 | 0.000 | 0.007 |
| 0.133 | 0.238 | 0.054 | 0.128 | 0.167 | 0.002 | 0.000 | 0.000 | 0.014 |
| 0.294 | 0.071 | 0.167 | 0.201 | 0.002 | 0.000 | 0.000 | 0.026 | 0.000 |
| 0.084 | 0.198 | 0.242 | 0.003 | 0.000 | 0.000 | 0.032 | 0.000 | 0.000 |
| 0.236 | 0.287 | 0.003 | 0.000 | 0.000 | 0.037 | 0.000 | 0.000 | 0.000 |
| 0.343 | 0.004 | 0.000 | 0.000 | 0.044 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.005 | 0.000 | 0.000 | 0.054 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.065 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 |
| 0.001 | 0.081 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 |
| 0.106 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.003 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.005 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.008 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.014 |
| 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.023 |
| 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.037 |
| 0.002 | 0.000 | 0.006 | 0.008 | 0.007 | 0.000 | 0.000 | 0.000 | 0.061 |
| 0.002 | 0.021 | 0.038 | 0.038 | 0.000 | 0.000 | 0.000 | 0.000 | 0.101 |
| 0.037 | 0.080 | 0.083 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.168 |
| 0.238 | 0.469 | 0.003 | 0.002 | 0.002 | 0.003 | 0.004 | 0.003 | 0.276 |
| 2.400 | 0.013 | 0.008 | 0.006 | 0.013 | 0.014 | 0.022 | 0.020 | 0.455 |
| 0.056 | 0.025 | 0.025 | 0.043 | 0.028 | 0.100 | 0.100 | 0.148 | 0.719 |
| 0.045 | 0.037 | 0.071 | 0.076 | 0.224 | 0.233 | 0.321 | 1.627 | 0.148 |
| 0.047 | 0.109 | 0.120 | 0.364 | 0.359 | 0.538 | 2.724 | 0.180 | 0.295 |
| 0.146 | 0.149 | 0.539 | 0.517 | 0.813 | 4.160 | 0.294 | 0.286 | 0.573 |


| 1962 | 19.003 | 26.983 | 19.058 | 20.474 | 0.170 | 0.078 | 0.075 | 0.192 | 0.186 | 0.690 | 0.658 | 1.070 | 5.710 | 0.390 | 0.391 | 0.517 | 0.589 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 76.103 | 72.088 | 54.883 | 0.231 | 0.099 | 0.092 | 0.239 | 0.237 | 0.868 | 0.817 | 1.390 | 7.419 | 0.506 | 0.503 | 0.669 | 0.233 | 0.849 |
| 1960 | 197.514 | 156.290 | 1.185 | 0.247 | 0.127 | 0.308 | 0.302 | 1.112 | 1.029 | 1.835 | 9.874 | 0.680 | 0.725 | 0.942 | 0.331 | 0.392 | 1.057 |
| 1959 | 412.478 | 3.225 | 1.117 | 0.163 | 0.387 | 0.379 | 1.449 | 1.319 | 2.391 | 12.661 | 0.886 | 0.976 | 1.308 | 0.480 | 0.539 | 0.860 | 1.040 |
| 1958 | 23.094 | 7.136 | 1.447 | 0.468 | 0.460 | 1.804 | 1.628 | 2.988 | 15.760 | 1.105 | 1.266 | 1.737 | 0.665 | 0.709 | 1.075 | 0.948 | 0.959 |
| 1957 | 25.397 | 8.719 | 1.496 | 0.559 | 2.498 | 2.058 | 3.718 | 20.450 | 1.362 | 1.608 | 2.214 | 0.867 | 0.955 | 1.410 | 1.244 | 0.381 | 1.440 |
| 1956 | 29.858 | 6.852 | 2.358 | 3.028 | 2.662 | 4.658 | 26.514 | 1.701 | 2.089 | 2.857 | 1.149 | 1.323 | 1.932 | 1.621 | 0.510 | 0.787 | 1.905 |
| 1955 | 24.971 | 10.303 | 8.247 | 3.193 | 5.710 | 33.009 | 2.099 | 2.631 | 3.616 | 1.427 | 1.745 | 2.564 | 2.102 | 0.656 | 1.005 | 1.124 | 2.629 |
| 1954 | 42.086 | 31.374 | 9.195 | 6.921 | 39.898 | 2.593 | 3.311 | 4.730 | 1.795 | 2.243 | 3.453 | 2.939 | 0.858 | 1.277 | 1.449 | 1.115 | 3.831 |
| 1953 | 86.102 | 30.543 | 17.934 | 47.153 | 3.063 | 3.956 | 5.878 | 2.174 | 2.726 | 4.351 | 3.840 | 1.063 | 1.581 | 1.785 | 1.402 | 3.082 | 4.103 |
| 1952 | 96.644 | 58.460 | 117.910 | 3.601 | 4.661 | 7.479 | 2.673 | 3.387 | 5.465 | 4.869 | 1.321 | 1.930 | 2.190 | 1.744 | 3.782 | 3.938 | 3.488 |
| 1951 | 146.355 | 301.944 | 9.484 | 5.422 | 9.103 | 3.292 | 4.113 | 6.905 | 6.306 | 1.617 | 2.331 | 2.656 | 2.142 | 4.667 | 4.950 | 0.757 | 5.388 |
| 1950 | 750.680 | 26.465 | 14.278 | 10.874 | 4.023 | 4.978 | 8.612 | 8.004 | 1.965 | 2.804 | 3.203 | 2.583 | 5.632 | 6.148 | 0.952 | 2.567 | 6.709 |

Table. NSSH. Modelled historic F-values.

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0.000 | 0.000 | 0.001 | 0.010 | 0.025 | 0.048 | 0.067 | 0.084 | 0.092 | 0.108 | 0.138 | 0.110 | 0.092 | 0.093 | 0.093 | 0.093 | 0.093 |
| 2004 | 0.000 | 0.000 | 0.001 | 0.024 | 0.047 | 0.071 | 0.094 | 0.185 | 0.223 | 0.273 | 0.324 | 0.374 | 0.273 | 0.264 | 0.273 | 0.262 | 1.340 |
| 2003 | 0.000 | 0.000 | 0.002 | 0.026 | 0.031 | 0.065 | 0.171 | 0.199 | 0.163 | 0.295 | 0.322 | 0.282 | 0.212 | 0.317 | 0.429 | 0.252 | 0.118 |
| 2002 | 0.000 | 0.000 | 0.013 | 0.016 | 0.048 | 0.182 | 0.269 | 0.156 | 0.245 | 0.259 | 0.227 | 0.233 | 0.259 | 0.358 | 0.542 | 0.241 | 0.281 |
| 2001 | 0.000 | 0.000 | 0.000 | 0.006 | 0.086 | 0.235 | 0.150 | 0.176 | 0.199 | 0.200 | 0.237 | 0.242 | 0.401 | 0.166 | 1.120 | 0.230 | 0.261 |
| 2000 | 0.000 | 0.000 | 0.001 | 0.037 | 0.213 | 0.102 | 0.149 | 0.192 | 0.216 | 0.318 | 0.434 | 0.620 | 0.476 | 1.590 | 8.330 | 0.219 | 0.687 |
| 1999 | 0.000 | 0.000 | 0.001 | 0.040 | 0.082 | 0.135 | 0.148 | 0.188 | 0.232 | 0.342 | 0.463 | 0.256 | 0.793 | 0.236 | 1.430 | 0.208 | 0.085 |
| 1998 | 0.000 | 0.000 | 0.014 | 0.125 | 0.179 | 0.097 | 0.150 | 0.177 | 0.282 | 0.330 | 0.450 | 0.261 | 0.084 | 0.606 | 6.820 | 0.198 | 0.000 |
| 1997 | 0.000 | 0.000 | 0.021 | 0.073 | 0.057 | 0.115 | 0.199 | 0.352 | 0.475 | 0.369 | 0.145 | 0.497 | 0.273 | 1.370 | 0.424 | 0.262 | 0.000 |
| 1996 | 0.000 | 0.000 | 0.010 | 0.006 | 0.037 | 0.115 | 0.274 | 0.340 | 0.373 | 0.033 | 0.074 | 0.140 | 0.463 | 0.514 | 0.001 | 0.303 | 0.000 |
| 1995 | 0.000 | 0.000 | 0.000 | 0.002 | 0.021 | 0.128 | 0.329 | 0.469 | 0.072 | 0.124 | 0.112 | 0.925 | 0.320 | 1.200 | 4.050 | 0.100 | 0.014 |
| 1994 | 0.000 | 0.000 | 0.000 | 0.002 | 0.018 | 0.129 | 0.207 | 0.059 | 0.050 | 0.048 | 0.200 | 0.154 | 0.349 | 0.939 | 2.220 | 0.100 | 0.000 |
| 1993 | 0.000 | 0.000 | 0.000 | 0.004 | 0.030 | 0.082 | 0.027 | 0.019 | 0.031 | 0.078 | 0.075 | 0.000 | 0.001 | 0.008 | 0.000 | 0.059 | 0.000 |
| 1992 | 0.000 | 0.000 | 0.000 | 0.003 | 0.025 | 0.013 | 0.005 | 0.011 | 0.020 | 0.034 | 0.175 | 0.486 | 0.963 | 0.041 | 0.048 | 0.029 | 0.000 |
| 1991 | 0.000 | 0.000 | 0.000 | 0.005 | 0.006 | 0.005 | 0.011 | 0.026 | 0.027 | 0.130 | 0.214 | 0.172 | 0.019 | 1.650 | 1.990 | 0.024 | 0.004 |
| 1990 | 0.000 | 0.000 | 0.006 | 0.035 | 0.008 | 0.008 | 0.026 | 0.024 | 0.053 | 0.436 | 1.430 | 0.056 | 1.680 | 0.413 | 0.022 | 0.153 | 0.266 |
| 1989 | 0.000 | 0.000 | 0.027 | 0.008 | 0.002 | 0.012 | 0.029 | 0.113 | 0.149 | 0.198 | 0.062 | 0.912 | 0.744 | 0.010 | 0.171 | 0.090 | 0.000 |
| 1988 | 0.001 | 0.001 | 0.015 | 0.030 | 0.043 | 0.040 | 0.223 | 0.436 | 0.857 | 0.206 | 1.460 | 1.060 | 0.088 | 0.875 | 4.760 | 0.381 | 0.000 |
| 1987 | 0.002 | 0.004 | 0.010 | 0.028 | 0.030 | 0.293 | 0.254 | 0.461 | 0.268 | 0.639 | 1.060 | 0.093 | 0.903 | 1.520 | 1.520 | 0.417 | 0.000 |
| 1986 | 0.004 | 0.000 | 0.003 | 0.027 | 0.187 | 0.588 | 0.530 | 0.565 | 1.260 | 1.250 | 0.835 | 0.820 | 2.060 | 2.160 | 0.001 | 1.400 | 0.594 |
| 1985 | 0.001 | 0.004 | 0.006 | 0.165 | 0.343 | 0.320 | 0.378 | 0.396 | 0.590 | 0.336 | 0.221 | 0.302 | 0.389 | 0.001 | 0.506 | 0.379 | 0.087 |
| 1984 | 0.005 | 0.000 | 0.010 | 0.070 | 0.074 | 0.121 | 0.083 | 0.085 | 0.059 | 0.116 | 0.112 | 0.035 | 0.000 | 0.294 | 0.242 | 0.070 | 0.000 |
| 1983 | 0.001 | 0.008 | 0.015 | 0.036 | 0.033 | 0.035 | 0.034 | 0.021 | 0.017 | 0.025 | 0.033 | 0.058 | 2.280 | 3.310 | 0.030 | 0.029 | 0.000 |
| 1982 | 0.015 | 0.004 | 0.001 | 0.018 | 0.025 | 0.021 | 0.017 | 0.022 | 0.023 | 0.023 | 0.040 | 0.539 | 0.194 | 0.004 | 0.024 | 0.022 | 0.000 |
| 1981 | 0.012 | 0.003 | 0.009 | 0.011 | 0.018 | 0.019 | 0.020 | 0.018 | 0.026 | 0.092 | 0.692 | 0.384 | 0.024 | 0.021 | 0.019 | 0.028 | 0.000 |
| 1980 | 0.007 | 0.000 | 0.001 | 0.021 | 0.011 | 0.018 | 0.027 | 0.042 | 0.093 | 0.033 | 0.002 | 0.055 | 0.017 | 0.016 | 0.023 | 0.058 | 0.000 |
| 1979 | 0.004 | 0.002 | 0.004 | 0.010 | 0.012 | 0.019 | 0.025 | 0.055 | 0.003 | 0.002 | 0.043 | 0.015 | 0.013 | 0.020 | 0.048 | 0.042 | 0.000 |
| 1978 | 0.005 | 0.002 | 0.001 | 0.017 | 0.028 | 0.037 | 0.115 | 0.003 | 0.743 | 0.069 | 0.012 | 0.011 | 0.017 | 0.039 | 0.035 | 0.079 | 0.000 |
| 1977 | 0.013 | 0.002 | 0.010 | 0.043 | 0.036 | 0.036 | 0.003 | 0.266 | 0.116 | 0.011 | 0.010 | 0.014 | 0.033 | 0.029 | 0.064 | 0.116 | 0.000 |
| 1976 | 0.003 | 0.003 | 0.001 | 0.030 | 0.397 | 0.002 | 0.000 | 0.109 | 0.009 | 0.008 | 0.012 | 0.027 | 0.024 | 0.052 | 0.090 | 0.015 | 0.000 |
| 1975 | 0.016 | 0.002 | 0.001 | 0.157 | 0.222 | 0.319 | 0.189 | 0.038 | 0.014 | 0.010 | 0.023 | 0.021 | 0.043 | 0.072 | 0.012 | 0.019 | 0.000 |
| 1974 | 0.012 | 0.002 | 0.097 | 0.122 | 0.060 | 0.112 | 0.973 | 0.779 | 0.009 | 0.019 | 0.017 | 0.035 | 0.058 | 0.011 | 0.016 | 0.017 | 0.000 |
| 1973 | 0.004 | 0.033 | 0.769 | 0.405 | 0.090 | 0.857 | 1.520 | 1.120 | 1.370 | 0.015 | 0.029 | 0.047 | 0.009 | 0.014 | 1.260 | 0.603 | 0.000 |
| 1972 | 0.842 | 1.830 | 0.987 | 0.098 | 1.290 | 0.956 | 1.820 | 1.250 | 2.980 | 1.780 | 0.039 | 1.570 | 2.070 | 1.110 | 0.332 | 2.110 | 0.000 |
| 1971 | 0.226 | 0.429 | 0.125 | 0.283 | 0.140 | 0.308 | 0.277 | 0.476 | 1.630 | 2.550 | 1.960 | 2.280 | 1.840 | 0.217 | 2.800 | 2.030 | 0.000 |
| 1970 | 0.333 | 0.261 | 1.300 | 0.529 | 1.490 | 0.298 | 0.678 | 1.100 | 1.160 | 1.180 | 1.370 | 1.550 | 3.940 | 1.620 | 0.823 | 1.410 | 0.024 |
| 1969 | 0.094 | 1.710 | 1.870 | 2.050 | 0.259 | 0.756 | 0.599 | 0.347 | 0.419 | 0.600 | 0.626 | 0.612 | 0.867 | 0.394 | 0.843 | 0.578 | 0.015 |
| 1968 | 0.775 | 0.722 | 0.724 | 3.230 | 4.590 | 4.750 | 1.810 | 1.210 | 0.936 | 1.580 | 1.290 | 1.320 | 0.771 | 1.960 | 1.250 | 1.360 | 0.067 |
| 1967 | 0.186 | 1.850 | 0.339 | 0.489 | 1.100 | 0.853 | 1.310 | 1.690 | 1.480 | 1.330 | 1.450 | 1.080 | 1.260 | 1.130 | 1.990 | 1.480 | 0.208 |
| 1966 | 0.128 | 0.742 | 0.244 | 0.308 | 0.405 | 0.529 | 0.593 | 0.749 | 1.310 | 0.976 | 1.340 | 1.060 | 0.543 | 1.370 | 1.480 | 1.190 | 0.280 |
| 1965 | 0.554 | 0.181 | 0.198 | 0.672 | 0.177 | 0.156 | 0.297 | 0.280 | 0.444 | 0.239 | 0.344 | 0.838 | 0.659 | 0.691 | 0.627 | 0.620 | 0.948 |
| 1964 | 0.062 | 0.067 | 0.536 | 0.061 | 0.083 | 0.187 | 0.138 | 0.031 | 0.072 | 0.283 | 0.306 | 0.333 | 0.284 | 0.366 | 0.365 | 0.320 | 0.849 |


| 1963 | 0.046 | 0.964 | 0.483 | 0.123 | 0.058 | 0.042 | 0.031 | 0.069 | 0.145 | 0.070 | 0.243 | 0.214 | 0.263 | 0.274 | 0.344 | 0.257 | 0.156 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1962 | 0.364 | 0.271 | 0.090 | 0.102 | 0.052 | 0.044 | 0.109 | 0.121 | 0.071 | 0.097 | 0.090 | 0.125 | 0.167 | 0.130 | 0.163 | 0.147 | 0.143 |
| 1961 | 0.137 | 0.430 | 0.086 | 0.157 | 0.092 | 0.049 | 0.070 | 0.092 | 0.080 | 0.067 | 0.112 | 0.112 | 0.112 | 0.101 | 0.107 | 0.106 | 0.040 |
| 1960 | 0.108 | 0.147 | 0.733 | 0.759 | 0.168 | 0.104 | 0.091 | 0.098 | 0.080 | 0.128 | 0.136 | 0.145 | 0.214 | 0.193 | 0.204 | 0.139 | 0.084 |
| 1959 | 0.071 | 0.102 | 0.611 | 0.105 | 0.078 | 0.077 | 0.115 | 0.099 | 0.115 | 0.099 | 0.114 | 0.148 | 0.178 | 0.221 | 0.168 | 0.113 | 0.101 |
| 1958 | 1.070 | 0.954 | 1.280 | 0.041 | 0.043 | 0.069 | 0.061 | 0.073 | 0.069 | 0.071 | 0.111 | 0.133 | 0.174 | 0.125 | 0.074 | 0.082 | 0.192 |
| 1957 | 0.369 | 0.896 | 0.262 | 0.046 | 0.176 | 0.084 | 0.069 | 0.111 | 0.059 | 0.089 | 0.093 | 0.116 | 0.147 | 0.121 | 0.121 | 0.100 | 0.151 |
| 1956 | 0.331 | 0.622 | 0.540 | 0.042 | 0.107 | 0.076 | 0.110 | 0.072 | 0.111 | 0.105 | 0.131 | 0.176 | 0.165 | 0.115 | 0.143 | 0.130 | 0.219 |
| 1955 | 0.393 | 0.575 | 0.102 | 0.032 | 0.054 | 0.069 | 0.061 | 0.081 | 0.085 | 0.067 | 0.127 | 0.133 | 0.110 | 0.101 | 0.095 | 0.104 | 0.175 |
| 1954 | 0.507 | 0.436 | 0.158 | 0.042 | 0.040 | 0.061 | 0.080 | 0.118 | 0.080 | 0.101 | 0.148 | 0.185 | 0.118 | 0.089 | 0.104 | 0.130 | 0.132 |
| 1953 | 0.110 | 0.301 | 0.052 | 0.017 | 0.017 | 0.028 | 0.067 | 0.041 | 0.045 | 0.081 | 0.117 | 0.065 | 0.064 | 0.059 | 0.079 | 0.078 | 0.085 |
| 1952 | 0.252 | 0.282 | 0.017 | 0.012 | 0.014 | 0.091 | 0.057 | 0.067 | 0.078 | 0.087 | 0.067 | 0.049 | 0.055 | 0.068 | 0.055 | 0.073 | 0.096 |
| 1951 | 0.018 | 0.040 | 0.068 | 0.001 | 0.047 | 0.058 | 0.044 | 0.084 | 0.109 | 0.053 | 0.039 | 0.043 | 0.056 | 0.060 | 0.079 | 0.070 | 0.050 |
| 1950 | 0.011 | 0.126 | 0.068 | 0.028 | 0.051 | 0.041 | 0.071 | 0.088 | 0.045 | 0.035 | 0.038 | 0.037 | 0.038 | 0.067 | 0.078 | 0.046 | 0.033 |



Figure 3.7.2.1.12. NSSH. Retrospective plot for the assessment in 2004.

Table 3.7.4.2 NSSH. Stock summary table. Biomass in million tonnes.

|  | Recruits | Total | Spawning stock | Fbar |
| :---: | :---: | :---: | :---: | :---: |
|  | Age 0 | biomass | biomass | 5-14 |
| 1950 | 750.680 | 20.013 | 14.359 | 0.058 |
| 1951 | 146.355 | 19.274 | 12.635 | 0.070 |
| 1952 | 96.644 | 20.182 | 11.042 | 0.073 |
| 1953 | 86.102 | 17.419 | 9.457 | 0.066 |
| 1954 | 42.086 | 18.565 | 8.703 | 0.113 |
| 1955 | 24.971 | 15.725 | 9.324 | 0.078 |
| 1956 | 29.858 | 13.799 | 10.934 | 0.110 |
| 1957 | 25.397 | 11.088 | 9.661 | 0.103 |
| 1958 | 23.094 | 9.549 | 8.731 | 0.079 |
| 1959 | 412.478 | 8.076 | 7.200 | 0.113 |
| 1960 | 197.514 | 7.634 | 5.853 | 0.136 |
| 1961 | 76.103 | 7.796 | 4.403 | 0.104 |
| 1962 | 19.003 | 6.765 | 3.443 | 0.146 |
| 1963 | 168.931 | 6.913 | 2.641 | 0.253 |
| 1964 | 93.903 | 6.446 | 2.479 | 0.226 |
| 1965 | 8.491 | 5.935 | 2.996 | 0.278 |
| 1966 | 51.409 | 4.392 | 2.658 | 0.696 |
| 1967 | 3.947 | 3.018 | 1.304 | 1.519 |
| 1968 | 5.187 | 0.982 | 0.318 | 3.493 |
| 1969 | 9.785 | 0.190 | 0.142 | 0.590 |
| 1970 | 0.661 | 0.116 | 0.069 | 1.320 |
| 1971 | 0.236 | 0.130 | 0.032 | 1.525 |
| 1972 | 0.957 | 0.085 | 0.016 | 1.497 |
| 1973 | 12.884 | 0.112 | 0.086 | 1.173 |
| 1974 | 8.631 | 0.160 | 0.091 | 0.114 |
| 1975 | 2.971 | 0.302 | 0.079 | 0.190 |
| 1976 | 10.068 | 0.362 | 0.139 | 0.106 |
| 1977 | 5.095 | 0.429 | 0.288 | 0.111 |
| 1978 | 6.201 | 0.579 | 0.360 | 0.043 |
| 1979 | 12.498 | 0.635 | 0.391 | 0.024 |
| 1980 | 1.474 | 0.748 | 0.475 | 0.034 |
| 1981 | 1.100 | 0.796 | 0.509 | 0.022 |
| 1982 | 2.343 | 0.729 | 0.507 | 0.020 |
| 1983 | 322.362 | 1.086 | 0.579 | 0.029 |
| 1984 | 11.528 | 2.009 | 0.603 | 0.090 |
| 1985 | 35.051 | 5.166 | 0.502 | 0.379 |
| 1986 | 6.041 | 1.815 | 0.401 | 1.074 |
| 1987 | 8.945 | 3.117 | 0.877 | 0.404 |
| 1988 | 25.009 | 3.485 | 2.738 | 0.045 |
| 1989 | 67.357 | 4.067 | 3.335 | 0.029 |
| 1990 | 114.598 | 4.550 | 3.490 | 0.022 |
| 1991 | 318.995 | 5.187 | 3.628 | 0.025 |
| 1992 | 371.421 | 6.208 | 3.496 | 0.029 |
| 1993 | 92.042 | 7.270 | 3.352 | 0.068 |
| 1994 | 29.993 | 8.275 | 3.775 | 0.139 |
| 1995 | 9.773 | 9.061 | 4.592 | 0.240 |
| 1996 | 57.485 | 9.123 | 6.113 | 0.197 |
| 1997 | 37.573 | 9.012 | 7.308 | 0.187 |
| 1998 | 258.857 | 7.841 | 6.564 | 0.168 |
| 1999 | 196.635 | 8.863 | 5.930 | 0.212 |
| 2000 | 45.323 | 7.986 | 4.635 | 0.262 |
| 2001 | 16.362 | 6.798 | 3.878 | 0.213 |
| 2002 | 288.899 | 7.220 | 3.918 | 0.232 |
| 2003 | 160.617 | 8.542 | 5.107 | 0.132 |
| 2004 | 401.287 | 10.180 | 6.513 | 0.119 |

Table 3.9.1.1 NSSH. Input to the short term prediction.

| Landings in 2005 |  |  |  | 1. million tonnes |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fbar age range: |  |  |  | 5-14, Fbar is weighted with population numbers January 1 |  |  |  |  |  |
| Total biomass in 2005 |  |  |  | 11.1161 million tonnes |  |  |  |  |  |
| Spawning stock biomass in 2005 |  |  |  | 6.13347 | million tonnes |  |  |  |  |
| Fbar in 2005 |  |  |  |  | 0.1846 |  |  |  |  |
|  | of year | spawni | 0.1 |  |  |  |  |  |  |
|  | Numbers | Weight | Weight | Weight | Weight | Fraction | Fraction | Fraction | loitation |
|  | (billion) | stock | stock | stock | catch | mature | atu | ature | ern |
|  | 2005 | 2005 | 2006 | 2007 | (kg) | 2005 | 2006 | 2007 |  |
| 0 | 0.000 | 0.001 | 0.001 | 0.001 | 0.007 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1 | 163.151 | 0.010 | 0.010 | 0.010 | 0.065 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 | 26.549 | 0.047 | 0.050 | 0.050 | 0.143 | 0.000 | 0.000 | 0.000 | 0.002 |
| 3 | 19.391 | 0.112 | 0.110 | 0.110 | 0.183 | 0.150 | 0.100 | 0.100 | 0.013 |
| 4 | 0.921 | 0.154 | 0.160 | 0.160 | 0.211 | 0.450 | 0.600 | 0.450 | 0.047 |
| 5 | 2.070 | 0.233 | 0.249 | 0.246 | 0.236 | 0.900 | 0.900 | 0.900 | 0.106 |
| 6 | 7.492 | 0.265 | 0.282 | 0.279 | 0.263 | 1.000 | 1.000 | 1.000 | 0.174 |
| 7 | 7.709 | 0.289 | 0.310 | 0.306 | 0.297 | 1.000 | 1.000 | 1.000 | 0.215 |
| 8 | 0.614 | 0.317 | 0.339 | 0.335 | 0.342 | 1.000 | 1.000 | 1.000 | 0.239 |
| 9 | 0.476 | 0.354 | 0.367 | 0.362 | 0.358 | 1.000 | 1.000 | 1.000 | 0.262 |
| 10 | 0.078 | 0.362 | 0.379 | 0.375 | 0.361 | 1.000 | 1.000 | 1.000 | 0.254 |
| 11 | 0.124 | 0.411 | 0.416 | 0.411 | 0.382 | 1.000 | 1.000 | 1.000 | 0.288 |
| 12 | 0.306 | 0.394 | 0.406 | 0.401 | 0.389 | 1.000 | 1.000 | 1.000 | 0.294 |
| 13 | 1.027 | 0.398 | 0.409 | 0.405 | 0.399 | 1.000 | 1.000 | 1.000 | 0.465 |
| 14 | 0.551 | 0.413 | 0.423 | 0.418 | 0.420 | 1.000 | 1.000 | 1.000 | 0.666 |
| 15 | 0.068 | 0.442 | 0.470 | 0.464 | 0.439 | 1.000 | 1.000 | 1.000 | 0.201 |
| 16 | 0.020 | 0.445 | 0.470 | 0.464 | 0.439 | 1.000 | 1.000 | 1.000 | 0.282 |

Table 3.9.1.2 NSSH. Short term prediction and two optional runs.
Short term prediction. Averaging period 10 Averaged

| SSB | Fbar Landings Biomass 3+ |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| 2006 |  |  |  | 2007 |
| 6.498 | 0.000 | 0.000 | 12.467 | 8.500 |
| 6.437 | 0.115 | 0.588 | 11.873 | 7.866 |
| 6.435 | 0.120 | 0.613 | 11.848 | 7.840 |
| 6.432 | 0.125 | 0.636 | 11.824 | 7.815 |
| 6.430 | 0.130 | 0.660 | 11.800 | 7.789 |
| 6.427 | 0.135 | 0.684 | 11.776 | 7.764 |
| 6.424 | 0.140 | 0.708 | 11.753 | 7.739 |
| 6.422 | 0.145 | 0.731 | 11.729 | 7.714 |
| 6.419 | 0.150 | 0.755 | 11.705 | 7.689 |
| 6.417 | 0.155 | 0.778 | 11.682 | 7.664 |
| 6.414 | 0.160 | 0.801 | 11.659 | 7.640 |

Optional 1. Short term prediction Averaging period 10 Averaged - Increase 1 year on 4 year olds
SSB Fbar Landings Biomass 3+ SSB
$2006 \quad 2006 \quad 2007 \quad 2007$
$\begin{array}{lllll}6.489 & 0.000 & 0.000 & 12.457 & 8.490\end{array}$
$\begin{array}{llllll}6.424 & 0.115 & 0.676 & 11.767 & 7.771\end{array}$
$\begin{array}{lllll}6.421 & 0.120 & 0.704 & 11.739 & 7.741\end{array}$
$\begin{array}{lllll}6.418 & 0.125 & 0.732 & 11.711 & 7.712\end{array}$
$\begin{array}{lllll}6.415 & 0.130 & 0.759 & 11.683 & 7.683\end{array}$
$\begin{array}{lllll}6.412 & 0.135 & 0.786 & 11.655 & 7.654\end{array}$
$\begin{array}{lllll}6.410 & 0.140 & 0.814 & 11.628 & 7.626\end{array}$
$\begin{array}{lllll}6.407 & 0.145 & 0.841 & 11.600 & 7.597\end{array}$
$\begin{array}{lllll}6.404 & 0.150 & 0.868 & 11.573 & 7.569\end{array}$
$\begin{array}{lllll}6.401 & 0.155 & 0.895 & 11.545 & 7.541\end{array}$
$\begin{array}{lllll}6.398 & 0.160 & 0.921 & 11.518 & 7.513\end{array}$
Optional 2. Catches in 2006 and SSB in 2007 if the fishing mortality is applied to spawning stock biomass

| Catch |  |  |
| :--- | :--- | :--- |
| 2006 | SSB |  |
| 2007 |  |  |
| 0.000 | 0.000 | 7.380 |
| 0.115 | 0.758 | 7.296 |
| 0.120 | 0.791 | 7.292 |
| 0.125 | 0.823 | 7.288 |
| 0.130 | 0.856 | 7.285 |
| 0.135 | 0.889 | 7.281 |
| 0.140 | 0.922 | 7.277 |
| 0.145 | 0.955 | 7.274 |
| 0.150 | 0.988 | 7.270 |
| 0.155 | 1.021 | 7.267 |
| 0.160 | 1.054 | 7.263 |

## Blue Whiting.

Table. Blue whiting. Stock numbers at age derived by AMCI.

| Age | $\mathbf{1 9 8 1}$ | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ |  |  |  |  |  |  |  |  |
| $\mathbf{1}$ | $3,634,926$ | $4,074,635$ | $10,886,489$ | $18,346,768$ | $12,241,432$ | $9,473,498$ | $8,982,704$ | $7,684,033$ |
| $\mathbf{2}$ | $3,994,375$ | $2,763,744$ | $3,179,573$ | $7,637,886$ | $12,598,817$ | $8,534,761$ | $6,686,965$ | $6,520,681$ |
| $\mathbf{3}$ | $4,787,469$ | $2,956,782$ | $2,090,780$ | $2,345,821$ | $5,324,109$ | $8,596,914$ | $5,530,846$ | $4,489,524$ |
| $\mathbf{4}$ | $3,032,825$ | $3,246,249$ | $2,105,565$ | $1,455,627$ | $1,547,086$ | $3,343,590$ | $4,952,787$ | $3,357,315$ |
| $\mathbf{5}$ | $2,421,173$ | $1,978,893$ | $2,236,277$ | $1,414,218$ | 921,378 | 952,453 | $1,798,949$ | $2,745,985$ |
| $\mathbf{6}$ | $2,309,452$ | $1,509,387$ | $1,327,705$ | $1,461,018$ | 863,349 | 546,013 | 496,258 | 968,288 |
| $\mathbf{7}$ | $1,878,814$ | $1,228,032$ | 903,621 | 759,007 | 748,554 | 417,017 | 222,861 | 222,973 |
| $\mathbf{8}$ | $1,799,023$ | 999,044 | 735,182 | 516,571 | 388,878 | 361,568 | 170,210 | 100,133 |
| $\mathbf{9}$ | $1,487,675$ | 956,615 | 598,095 | 420,281 | 264,666 | 187,837 | 147,578 | 76,477 |
| $\mathbf{1 0 +}$ | $3,149,132$ | $2,465,584$ | $2,048,759$ | $1,513,122$ | 990,581 | 606,312 | 324,140 | 211,947 |


| Age | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ |  |  |  |  |  |  |  |  |
| $\mathbf{1}$ | $8,938,636$ | $22,336,700$ | $9,008,614$ | $6,160,529$ | $5,525,904$ | $6,505,224$ | $7,617,300$ | $20,406,428$ |
| $\mathbf{2}$ | $5,786,361$ | $6,601,063$ | $16,718,297$ | $7,119,122$ | $4,769,354$ | $4,291,179$ | $5,060,322$ | $5,919,291$ |
| $\mathbf{3}$ | $4,492,794$ | $3,896,371$ | $4,504,638$ | $12,659,255$ | $5,362,474$ | $3,612,442$ | $3,279,957$ | $3,817,265$ |
| $\mathbf{4}$ | $2,841,816$ | $2,704,454$ | $2,391,453$ | $3,252,181$ | $9,078,501$ | $3,872,129$ | $2,628,043$ | $2,329,541$ |
| $\mathbf{5}$ | $1,958,331$ | $1,572,766$ | $1,520,596$ | $1,665,696$ | $2,242,592$ | $6,296,292$ | $2,722,504$ | $1,788,218$ |
| $\mathbf{6}$ | $1,505,873$ | 992,705 | 812,496 | $1,010,437$ | $1,105,252$ | $1,503,193$ | $4,251,225$ | $1,761,823$ |
| $\mathbf{7}$ | 444,766 | 626,251 | 412,484 | 490,239 | 610,836 | 685,201 | 950,396 | $2,574,480$ |
| $\mathbf{8}$ | 102,419 | 184,966 | 260,217 | 248,883 | 296,363 | 378,688 | 433,219 | 575,546 |
| $\mathbf{9}$ | 45,994 | 42,593 | 76,856 | 157,008 | 150,456 | 183,730 | 239,426 | 262,351 |
| $\mathbf{1 0 +}$ | 132,482 | 74,224 | 48,539 | 75,660 | 140,654 | 180,474 | 230,268 | 284,440 |


| Age | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathbf{0}$ |  |  |  |  |  |  |  |  |
| $\mathbf{1}$ | $40,137,768$ | $30,806,560$ | $19,527,742$ | $40,592,636$ | $64,009,280$ | $34,671,620$ | $41,366,872$ | $41,950,092$ |
| $\mathbf{2}$ | $15,314,992$ | $30,492,660$ | $23,208,058$ | $15,039,314$ | $31,336,016$ | $48,849,068$ | $26,757,696$ | $31,224,204$ |
| $\mathbf{3}$ | $4,353,480$ | $11,222,027$ | $21,317,802$ | $16,517,499$ | $10,575,065$ | $21,754,324$ | $34,914,972$ | $18,904,828$ |
| $\mathbf{4}$ | $2,589,741$ | $2,919,289$ | $6,837,649$ | $13,164,899$ | $9,881,424$ | $6,154,365$ | $13,411,988$ | $21,161,342$ |
| $\mathbf{5}$ | $1,489,306$ | $1,639,443$ | $1,640,733$ | $3,810,114$ | $6,916,746$ | $4,981,085$ | $3,292,873$ | $6,970,656$ |
| $\mathbf{6}$ | $1,083,852$ | 899,824 | 882,762 | 915,044 | $1,995,481$ | $3,480,721$ | $2,660,460$ | $1,639,702$ |
| $\mathbf{7}$ | 972,908 | 597,449 | 422,277 | 439,253 | 414,796 | 881,843 | $1,649,291$ | $1,216,932$ |
| $\mathbf{8}$ | $1,421,670$ | 536,294 | 280,376 | 210,121 | 199,116 | 183,306 | 417,849 | 754,409 |
| $\mathbf{9}$ | 317,826 | 783,664 | 251,677 | 139,512 | 95,249 | 87,993 | 86,857 | 191,130 |
| $\mathbf{1 0 +}$ | 301,947 | 341,636 | 528,091 | 388,004 | 239,127 | 147,767 | 111,712 | 90,828 |

Table. Blue whiting. Fishing mortality derived by AMCI.

| Age | $\mathbf{1 9 8 1}$ | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{0}$ |  |  |  |  |  |  |  |  |
| $\mathbf{1}$ | 0.074 | 0.048 | 0.154 | 0.176 | 0.161 | 0.148 | 0.120 | 0.084 |
| $\mathbf{2}$ | 0.101 | 0.079 | 0.104 | 0.161 | 0.182 | 0.234 | 0.198 | 0.173 |
| $\mathbf{3}$ | 0.189 | 0.140 | 0.162 | 0.216 | 0.265 | 0.351 | 0.299 | 0.257 |
| $\mathbf{4}$ | 0.227 | 0.173 | 0.198 | 0.257 | 0.285 | 0.420 | 0.390 | 0.339 |
| $\mathbf{5}$ | 0.273 | 0.199 | 0.226 | 0.294 | 0.323 | 0.452 | 0.419 | 0.401 |
| $\mathbf{6}$ | 0.432 | 0.313 | 0.359 | 0.469 | 0.528 | 0.696 | 0.600 | 0.578 |
| $\mathbf{7}$ | 0.432 | 0.313 | 0.359 | 0.469 | 0.528 | 0.696 | 0.600 | 0.578 |
| $\mathbf{8}$ | 0.432 | 0.313 | 0.359 | 0.469 | 0.528 | 0.696 | 0.600 | 0.578 |
| $\mathbf{9}$ | 0.432 | 0.313 | 0.359 | 0.469 | 0.528 | 0.696 | 0.600 | 0.578 |
| $\mathbf{1 0 +}$ | 0.432 | 0.313 | 0.359 | 0.469 | 0.528 | 0.696 | 0.600 | 0.578 |
|  |  |  |  |  |  |  |  |  |
| F(3-7) | 0.310 | 0.227 | 0.261 | 0.341 | 0.386 | 0.523 | 0.462 | 0.431 |
|  |  |  |  |  |  |  |  |  |
| Age | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ |
| $\mathbf{0}$ |  |  |  |  |  |  |  |  |
| $\mathbf{1}$ | 0.103 | 0.090 | 0.035 | 0.056 | 0.053 | 0.051 | 0.052 | 0.087 |
| $\mathbf{2}$ | 0.195 | 0.182 | 0.078 | 0.083 | 0.078 | 0.069 | 0.082 | 0.107 |
| $\mathbf{3}$ | 0.308 | 0.288 | 0.126 | 0.132 | 0.126 | 0.118 | 0.142 | 0.188 |
| $\mathbf{4}$ | 0.392 | 0.376 | 0.162 | 0.172 | 0.166 | 0.152 | 0.185 | 0.247 |
| $\mathbf{5}$ | 0.479 | 0.460 | 0.209 | 0.210 | 0.200 | 0.193 | 0.235 | 0.301 |
| $\mathbf{6}$ | 0.677 | 0.678 | 0.305 | 0.303 | 0.278 | 0.258 | 0.302 | 0.394 |
| $\mathbf{7}$ | 0.677 | 0.678 | 0.305 | 0.303 | 0.278 | 0.258 | 0.302 | 0.394 |
| $\mathbf{8}$ | 0.677 | 0.678 | 0.305 | 0.303 | 0.278 | 0.258 | 0.302 | 0.394 |
| $\mathbf{9}$ | 0.677 | 0.678 | 0.305 | 0.303 | 0.278 | 0.258 | 0.302 | 0.394 |
| $\mathbf{1 0 +}$ | 0.677 | 0.678 | 0.305 | 0.303 | 0.278 | 0.258 | 0.302 | 0.394 |
|  |  |  |  |  |  |  |  |  |
| F(3-7) | 0.507 | 0.496 | 0.221 | 0.224 | 0.210 | 0.196 | 0.233 | 0.305 |
|  |  |  |  |  |  |  |  |  |
| Age | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ |
| $\mathbf{0}$ |  |  |  |  |  |  |  |  |
| $\mathbf{1}$ | 0.075 | 0.083 | 0.061 | 0.059 | 0.070 | 0.059 | 0.081 | 0.073 |
| $\mathbf{2}$ | 0.111 | 0.158 | 0.140 | 0.152 | 0.165 | 0.136 | 0.147 | 0.167 |
| $\mathbf{3}$ | 0.200 | 0.295 | 0.282 | 0.314 | 0.341 | 0.284 | 0.301 | 0.352 |
| $\mathbf{4}$ | 0.257 | 0.376 | 0.385 | 0.444 | 0.485 | 0.425 | 0.454 | 0.505 |
| $\mathbf{5}$ | 0.304 | 0.419 | 0.384 | 0.447 | 0.487 | 0.427 | 0.497 | 0.568 |
| $\mathbf{6}$ | 0.396 | 0.557 | 0.498 | 0.591 | 0.617 | 0.547 | 0.582 | 0.719 |
| $\mathbf{7}$ | 0.396 | 0.557 | 0.498 | 0.591 | 0.617 | 0.547 | 0.582 | 0.719 |
| $\mathbf{8}$ | 0.396 | 0.557 | 0.498 | 0.591 | 0.617 | 0.547 | 0.582 | 0.719 |
| $\mathbf{9}$ | 0.396 | 0.557 | 0.498 | 0.591 | 0.617 | 0.547 | 0.582 | 0.719 |
| $\mathbf{1 0 +}$ | 0.396 | 0.557 | 0.498 | 0.591 | 0.617 | 0.547 | 0.582 | 0.719 |
|  |  |  |  |  |  |  |  |  |
| F(3-7) | 0.310 | 0.441 | 0.409 | 0.477 | 0.509 | 0.446 | 0.483 | 0.572 |
|  |  |  |  |  |  |  |  |  |

Table. Blue whiting. Summary table by AMCI.
Run id 20050829194934.741

## SUMMARY TABLE

| Year | Recruits <br> age 1 | SSB <br> $3-7$ | F Catch <br> SOP |  |
| :---: | :---: | :---: | :---: | :---: |
| 1981 | 3634925 | 2840186 | 0.3102 | 922980 |
| 1982 | 4074635 | 2340060 | 0.2275 | 550643 |
| 1983 | 10886489 | 1903785 | 0.2608 | 553344 |
| 1984 | 18346767 | 1548446 | 0.3409 | 615569 |
| 1985 | 12241432 | 1651896 | 0.3858 | 678214 |
| 1986 | 9473497 | 1855547 | 0.5231 | 847145 |
| 1987 | 8982703 | 1655472 | 0.4617 | 654718 |
| 1988 | 7684032 | 1469257 | 0.4306 | 552264 |
| 1989 | 8938635 | 1412695 | 0.5067 | 630316 |
| 1990 | 22336699 | 1335867 | 0.4962 | 558128 |
| 1991 | 9008614 | 1803269 | 0.2213 | 364008 |
| 1992 | 6160528 | 2437900 | 0.2242 | 474592 |
| 1993 | 5525903 | 2385790 | 0.2096 | 475198 |
| 1994 | 6505223 | 2354832 | 0.1960 | 457696 |
| 1995 | 7617300 | 2180177 | 0.2331 | 505175 |
| 1996 | 20406428 | 2019352 | 0.3047 | 621104 |
| 1997 | 40137767 | 2062619 | 0.3104 | 639680 |
| 1998 | 30806559 | 2827013 | 0.4408 | 1131954 |
| 1999 | 19527741 | 3434783 | 0.4093 | 1261033 |
| 2000 | 40592635 | 3562036 | 0.4773 | 1412449 |
| 2001 | 64009279 | 4005004 | 0.5093 | 1771805 |
| 2002 | 34671620 | 4881275 | 0.4460 | 1556954 |
| 2003 | 41366871 | 5729501 | 0.4834 | 2365319 |
| 2004 | 41950091 | 5113236 | 0.5725 | 2383503 |

Table. Blue Whiting. Input data for short term prediction.
MFDP version 1
Run: AMCI run 2
Time and date: 15:33 29/09/2005
Fbar age range: 3-7
Recruitment, geometric mean of 1996-2003.

| 2005 |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | N |  | M |  | Mat |  | PF | PM |  | SWt | Sel | CWt |
| 1 |  | 34118628 |  | 0.2 |  | 0.11 | 0.25 |  | 0.25 | 0.049 | 0.081 | 0.049 |
| 2 |  | 31918788 |  | 0.2 |  | 0.4 | 0.25 |  | 0.25 | 0.073 | 0.175 | 0.073 |
| 3 |  | 21639784 |  | 0.2 |  | 0.82 | 0.25 |  | 0.25 | 0.094 | 0.364 | 0.094 |
| 4 |  | 10888237 |  | 0.2 |  | 0.86 | 0.25 |  | 0.25 | 0.108 | 0.532 | 0.108 |
| 5 |  | 10452690 |  | 0.2 |  | 0.91 | 0.25 |  | 0.25 | 0.129 | 0.563 | 0.129 |
| 6 |  | 3234738 |  | 0.2 |  | 0.94 | 0.25 |  | 0.25 | 0.153 | 0.702 | 0.153 |
| 7 |  | 654252.9 |  | 0.2 |  | 1 | 0.25 |  | 0.25 | 0.172 | 0.702 | 0.172 |
| 8 |  | 485564.7 |  | 0.2 |  | 1 | 0.25 |  | 0.25 | 0.198 | 0.702 | 0.198 |
| 9 |  | 301014.8 |  | 0.2 |  | 1 | 0.25 |  | 0.25 | 0.224 | 0.702 | 0.224 |
| 10 |  | 112503.5 |  | 0.2 |  | 1 | 0.25 |  | 0.25 | 0.278 | 0.702 | 0.278 |

2006

| Age | N |  | M |  | Mat |  | PF |  | PM |  | SWt | Sel |  | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  | 34118628 |  | 0.2 |  | 0.11 |  | 0.25 |  | 0.25 | 0.049 |  | 0.081 | 0.049 |
| 2 | . |  |  | 0.2 |  | 0.4 |  | 0.25 |  | 0.25 | 0.073 |  | 0.175 | 0.073 |
| 3 | . |  |  | 0.2 |  | 0.82 |  | 0.25 |  | 0.25 | 0.094 |  | 0.364 | 0.094 |
| 4 | . |  |  | 0.2 |  | 0.86 |  | 0.25 |  | 0.25 | 0.108 |  | 0.532 | 0.108 |
| 5 | . |  |  | 0.2 |  | 0.91 |  | 0.25 |  | 0.25 | 0.129 |  | 0.563 | 0.129 |
| 6 | . |  |  | 0.2 |  | 0.94 |  | 0.25 |  | 0.25 | 0.153 |  | 0.702 | 0.153 |
| 7 | . |  |  | 0.2 |  | 1 |  | 0.25 |  | 0.25 | 0.172 |  | 0.702 | 0.172 |
| 8 | . |  |  | 0.2 |  | 1 |  | 0.25 |  | 0.25 | 0.198 |  | 0.702 | 0.198 |
| 9 | - |  |  | 0.2 |  | 1 |  | 0.25 |  | 0.25 | 0.224 |  | 0.702 | 0.224 |
| 10 | . |  |  | 0.2 |  | 1 |  | 0.25 |  | 0.25 | 0.278 |  | 0.702 | 0.278 |



Input units are thousands and kg - output in tonnes

Table. Blue Whiting. Short term prediction.

| 2005 |  |  |  |  |
| ---: | :--- | ---: | ---: | ---: |
| Biomass | SSB | FMult | FBar | Landings |
| 9351161 | 4935621 |  | 1 | 0.5725 |
| 2366726 |  |  |  |  |

20062007

| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| :--- | :--- | ---: | :--- | ---: | ---: | :--- | :--- |
| 8723133 | 5183464 | 0 | 0 | 0 | 10609585 | 6771015 |
| . | 5150761 | 0.056 | 0.0321 | 154401 | 10443755 | 6586761 |
| . | 5118297 | 0.112 | 0.0641 | 304857 | 10282225 | 6408829 |
| . | 5086070 | 0.168 | 0.0962 | 451485 | 10124864 | 6236976 |
| . | 5054078 | 0.224 | 0.1282 | 594399 | 9971550 | 6070970 |
| . | 5022320 | 0.28 | 0.1603 | 733709 | 9822159 | 5910590 |
| . | 4990793 | 0.336 | 0.1924 | 869522 | 9676575 | 5755620 |
| . | 4959497 | 0.392 | 0.2244 | 1001940 | 9534685 | 5605857 |
| . | 4928428 | 0.448 | 0.2565 | 1131064 | 9396379 | 5461104 |
| . | 4897585 | 0.504 | 0.2885 | 1256991 | 9261550 | 5321172 |
| . | 4866967 | 0.56 | 0.3206 | 1379813 | 9130096 | 5185880 |
| . | 4836571 | 0.616 | 0.3526 | 1499622 | 9001916 | 5055055 |
| . | 4806396 | 0.672 | 0.3847 | 1616505 | 8876915 | 4928529 |
| . | 4776440 | 0.728 | 0.4168 | 1730547 | 8755000 | 4806142 |
| . | 4746701 | 0.784 | 0.4488 | 1841831 | 8636080 | 4687740 |
| . | 4717178 | 0.84 | 0.4809 | 1950436 | 8520067 | 4573176 |
| . | 4687869 | 0.896 | 0.5129 | 2056440 | 8406877 | 4462308 |
| . | 4658772 | 0.952 | 0.545 | 2159916 | 8296427 | 4355000 |
| . | 4629885 | 1.008 | 0.5771 | 2260939 | 8188639 | 4251122 |
| . | 4601208 | 1.064 | 0.6091 | 2359576 | 8083436 | 4150546 |
| . | 4572738 | 1.12 | 0.6412 | 2455898 | 7980743 | 4053154 |

Input units are thousands and kg - output units are tonnes.

## Appendix

## Norwegian spring spawning herring.

Rescaled catch in numbers for 2004.

Age |  | Numbers | Mean weight |
| ---: | ---: | ---: |
| 0 | 0.000 | 0.022 |
| 1 | 0.002 | 0.065 |
| 2 | 0.037 | 0.141 |
| 3 | 0.024 | 0.181 |
| 4 | 0.108 | 0.211 |
| 5 | 0.591 | 0.240 |
| 6 | 0.818 | 0.271 |
| 7 | 0.135 | 0.303 |
| 8 | 0.129 | 0.347 |
| 9 | 0.027 | 0.359 |
| 10 | 0.052 | 0.368 |
| 11 | 0.151 | 0.387 |
| 12 | 0.350 | 0.392 |
| 13 | 0.180 | 0.400 |
| 14 | 0.023 | 0.423 |
| $15+$ | 0.010 | 0.438 |

Table. 3.7.2.1.1 NSSH. Summary of SeaStar exploratory runs.
Spawning stock
Juvenile stock
terminalF83
terminalF90
terminalF91
terminalF92
terminalF93
terminalF96
terminalF97
terminalF98
terminalF99
terminalF02
terminalF03
terminalF04
cat1
cat2
cat3
cat5
catLarvae2
cat4
cat6
catZero
distributionParameter1
distributionParameterLarvae
distributionParameter2
distParZero
Log-likelihood per term,
survey 1
Log-likelihood per term,
survey 2
Log-likelihood per term,
survey 3
Log-likelihood per term,
survey 4
Log-likelihood per term,
survey 5
Log-likelihood per term,
survey 6
Number of data points,
surve

| Default | EstimateM | IncreasedMPlusgroup |
| :---: | :---: | :---: |
| 6.102 | 5.141 | 6.133 |
| 5.006 | 4.364 | 4.983 |
| 0.198 | 0.217 | 0.198 |
| 0.280 | 0.381 | 0.273 |
| 0.268 | 0.334 | 0.264 |
| 0.278 | 0.353 | 0.273 |
| 0.377 | 0.439 | 0.374 |
| 0.226 | 0.279 | 0.223 |
| 0.187 | 0.223 | 0.185 |
| 0.095 | 0.115 | 0.094 |
| 0.072 | 0.087 | 0.070 |
| 0.001 | 0.001 | 0.001 |
| 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 |
| 1.053 | 1.257 | 1.057 |
| 0.846 | 1.012 | 0.831 |
| 0.943 | 1.119 | 0.944 |
| 1.143 | 1.379 | 1.133 |
| 4.699 | 5.349 | 4.841 |
| 0.345 | 0.367 | 0.340 |
| 0.494 | 0.571 | 0.485 |
| 0.003 | 0.002 | 0.003 |
| 0.486 | 0.478 | 0.491 |
| 0.674 | 0.764 | 0.627 |
| 1.945 | 1.822 | 1.210 |
| 0.249 | 0.250 | 0.251 |
|  |  |  |
| -1.506 | -1.494 | -1.505 |
|  |  |  |
| -1.238 | -1.234 | -1.237 |
|  |  |  |
| -1.823 | -1.813 | -1.827 |
|  |  |  |
| -3.453 | -3.432 | -3.280 |
|  |  |  |
| -1.523 | -1.487 | -1.526 |
|  |  |  |
| -3.797 | -3.766 | -3.643 |
|  |  |  |
|  |  | 27 |

TimeTrendLarvae N

| TimeTrendLarvae |
| :---: |
| 5.678 |
| 4.701 |
| 0.209 |
| 0.300 |
| 0.288 |
| 0.300 |
| 0.398 |
| 0.238 |
| 0.195 |
| 0.103 |
| 0.077 |
| 0.001 |
| 0.000 |
| 0.000 |
| 1.104 |
| 0.861 |
| 0.957 |
| 1.187 |
| 5.173 |
| 0.374 |
| 0.543 |
| 0.003 |
| 0.494 |
| 0.430 |
| 1.227 |
| 0.243 |
|  |
| -1.536 |
|  |
| -1.224 |
|  |
| -1.832 |
| -3.277 |
| -1.520 |
| -3.652 |

NoLarvae
5.411


| $\log$ | 200 |
| :--- | :--- |
| 6.120 |  |
| 5.111 |  |


2002
6.
5.
6.1
5.1

120
.111
CorrectingAgeBias

Number of data points,
survey 1
Number of data points,
survey 2
Survey 2
Number of data points,
$27 \quad 27$
$58 \quad 58$

27
5.411
4.633
0.208
0.208
0.312
0.312
0.295
0.295
0.308
0.308
0.404
0.404
0.246
0.246
0.200
0.108
0.081
0.001
0.000
0.530
0.285
0.274
0.274
0.448
0.448
0.242
0.187
0.095
0.070
0.0701
0.000
0.000
1.113
1.113
0.885
0.960
1.208
0.364
0.000
0.000
0.540
0.340
0.003
0.485
1.948
1.03
0.75
0.93
1.072
4.748
0.348
0.487
0.3487
0.003
0.525
0.238
0.67
1.20
0.25
$-1.541$
$-1.560$

$$
-1.246
$$

$-1.822$
$-3.452$
$-3.276$
$-1.513$

$$
-1.546
$$

$-3.629$
6.329
5.125
CompatibleISVPA
4.564
5.125
0.167
0.285

Compatib
4.564
.318
0.257
0.257
0.290

$$
\begin{aligned}
& 0.200 \\
& 0.278 \\
& 0.278
\end{aligned}
$$

0.290
0.307
0.217

$$
0.278
$$

$$
\begin{aligned}
& 0.278 \\
& 0.377
\end{aligned}
$$

$$
\begin{aligned}
& 0.377 \\
& 0.226
\end{aligned}
$$

$$
\begin{aligned}
& 0.226 \\
& 0.187
\end{aligned}
$$

$$
\begin{aligned}
& 0.187 \\
& 0.095
\end{aligned}
$$

$$
\begin{aligned}
& 0.095 \\
& 0.072
\end{aligned}
$$

$$
\begin{aligned}
& 0.072 \\
& 0.001
\end{aligned}
$$

$$
\begin{aligned}
& 0.001 \\
& 0.000
\end{aligned}
$$

$$
\begin{aligned}
& 0.000 \\
& 0.000
\end{aligned}
$$

$$
\begin{aligned}
& 0.000 \\
& 1.053
\end{aligned}
$$

$$
\begin{aligned}
& 1.053 \\
& 0.846
\end{aligned}
$$

$$
\begin{aligned}
& 0.846 \\
& 0.943
\end{aligned}
$$

$$
\begin{aligned}
& 0.943 \\
& 1.143
\end{aligned}
$$

$$
\begin{aligned}
& 1.143 \\
& 4.699
\end{aligned}
$$

$$
\begin{aligned}
& 4.699 \\
& 0.345
\end{aligned}
$$

$$
\begin{aligned}
& 0.345 \\
& 0.494
\end{aligned}
$$

$$
\begin{aligned}
& 0.494 \\
& 0.003
\end{aligned}
$$

$$
\begin{aligned}
& 0.003 \\
& 0.486 \\
& 0.671
\end{aligned}
$$

$$
\begin{aligned}
& 0.486 \\
& 0.674 \\
& 1.945
\end{aligned}
$$

$$
\begin{aligned}
& 1.945 \\
& 0.249
\end{aligned}
$$

$$
-1.506
$$

$$
-1.238
$$

$$
-1.835
$$

$$
-3.457
$$

$$
-3.453
$$

$$
-1.499
$$

$$
-1.507
$$

| survey 3 <br> Number of data points, <br> survey 4 | 18 | 18 | 18 |
| :--- | ---: | ---: | ---: |
| Number of data points, <br> survey 5 | 20 | 20 | 20 |
| Number of data points, | 65 | 65 | 65 |
| survey 6 |  |  |  |
| vpaM | 11 | 11 | 11 |
| vpaMYoung | 0.150 | 0.104 | 0.150 |
| survLogLik1 | 0.900 | 1.077 | 0.900 |
| survLogLik2 | -244.240 | -241.169 | -244.424 |
| tagLogLik1 | -110.823 | -110.061 | -105.663 |
| tagLogLik2 | -369.756 | -365.624 | -369.722 |
| larvLogLik1 | 0.000 | 0.000 | 0.000 |
| larvLogLik2 | -80.560 | -84.084 | -78.634 |
| zeroLogLik1 | 0.000 | 0.000 | 0.000 |
| zeroLogLik2 | 0.000 | 0.000 | 0.000 |
| predLogLik1 | -0.362 | -0.104 | -0.460 |
| predLogLik2 | 0.000 | 0.000 | 0.000 |
|  | 0.000 | 0.000 | 0.000 |


| 18 | 18 | 18 | 0 |
| ---: | ---: | ---: | ---: |
| 20 |  |  |  |
|  | 20 | 20 | 40 |
| 65 |  |  | 65 |
|  |  |  | 0 |
| 11 | 11 | 11 | 22 |
| 0.150 | 0.150 | 0.150 | 0.150 |
| 0.900 | 0.900 | 0.900 | 0.900 |
| -244.198 | -243.987 | -248.199 | -105.438 |
| -105.716 | -110.879 | -105.438 | -105.438 |
| -369.545 | -369.620 | -369.093 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 |
| -68.730 | 0.000 | -80.624 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | -0.870 |
| -0.075 | 0.215 | -0.870 | -0.870 |
| 0.000 | 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 | 0.000 |


| 18 | 0 | 18 |
| ---: | :---: | ---: |
| 20 | 0 | 20 |
|  |  |  |
| 65 | 65 | 65 |
|  |  | 11 |
| 0.150 | 0.150 | 0.150 |
| 0.900 | 0.900 | 0.900 |
| -243.357 | -167.040 | -244.240 |
| -110.919 | 0.000 | -110.823 |
| 0.000 | 0.000 | -369.756 |
| 0.000 | 0.000 | 0.000 |
| -80.387 | 0.000 | -80.560 |
| 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 |
| -0.432 | 0.678 | -0.362 |
| 0.000 | 0.000 | 0.000 |
| 0.000 | 0.000 | 0.000 |

Table 3.7.4.1. NSSH. Stock in numbers.
2005
2004
2003
2002
2001
2000
1999
1998
1997
1996
1995
1994
1993
1992
1991
1990
1989
1988
1987
1986
1985
1984
1983
1982
1981
1980
1979
1978
1977
1976
1975
1974
1973
1972
1971
1970
1969
1968
1967
1966
1965
1964
1963

|  |  |  |
| ---: | ---: | ---: |
| 0 | 1 |  |
| 0.000 | 163.151 | 26.549 |
| 401.287 | 65.302 | 47.753 |
| 160.617 | 117.458 | 2.705 |
| 288.899 | 6.652 | 7.492 |
| 16.362 | 18.427 | 32.504 |
| 45.323 | 79.946 | 42.789 |
| 196.635 | 105.244 | 6.211 |
| 258.857 | 15.276 | 9.502 |
| 37.573 | 23.372 | 1.616 |
| 57.485 | 3.974 | 4.958 |
| 9.773 | 12.194 | 15.213 |
| 29.993 | 37.417 | 61.395 |
| 92.042 | 151.007 | 52.729 |
| 371.421 | 129.694 | 18.941 |
| 318.995 | 46.592 | 11.132 |
| 114.598 | 27.381 | 4.129 |
| 67.357 | 10.158 | 1.473 |
| 25.009 | 3.628 | 0.991 |
| 8.945 | 2.447 | 5.786 |
| 6.041 | 14.232 | 1.888 |
| 35.051 | 4.665 | 53.252 |
| 11.528 | 130.982 | 0.378 |
| 322.362 | 0.938 | 0.179 |
| 2.343 | 0.442 | 0.241 |
| 1.100 | 0.595 | 2.057 |
| 1.474 | 5.060 | 1.017 |
| 12.498 | 2.508 | 0.830 |
| 6.201 | 2.044 | 1.655 |
| 5.095 | 4.080 | 0.482 |
| 10.068 | 1.188 | 1.407 |
| 2.971 | 3.467 | 2.117 |
| 8.631 | 5.220 | 0.066 |
| 12.884 | 0.168 | 0.005 |
| 0.957 | 0.077 | 0.051 |
| 0.236 | 0.193 | 1.134 |
| 0.661 | 3.620 | 0.072 |
| 9.785 | 0.972 | 0.263 |
| 5.187 | 1.333 | 1.182 |
| 3.947 | 18.398 | 0.384 |
| 51.409 | 1.984 | 12.175 |
| 8.491 | 35.874 | 24.938 |
| 93.903 | 65.617 | 0.833 |
| 168.931 | 5.371 | 8.368 |
|  |  |  |


| 1962 | 19.003 | 26.983 | 19.058 | 20.474 | 0.170 | 0.078 | 0.075 | 0.192 | 0.186 | 0.690 | 0.658 | 1.070 | 5.710 | 0.390 | 0.391 | 0.517 | 0.589 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 76.103 | 72.088 | 54.883 | 0.231 | 0.099 | 0.092 | 0.239 | 0.237 | 0.868 | 0.817 | 1.390 | 7.419 | 0.506 | 0.503 | 0.669 | 0.233 | 0.849 |
| 1960 | 197.514 | 156.290 | 1.185 | 0.247 | 0.127 | 0.308 | 0.302 | 1.112 | 1.029 | 1.835 | 9.874 | 0.680 | 0.725 | 0.942 | 0.331 | 0.392 | 1.057 |
| 1959 | 412.478 | 3.225 | 1.117 | 0.163 | 0.387 | 0.379 | 1.449 | 1.319 | 2.391 | 12.661 | 0.886 | 0.976 | 1.308 | 0.480 | 0.539 | 0.860 | 1.040 |
| 1958 | 23.094 | 7.136 | 1.447 | 0.468 | 0.460 | 1.804 | 1.628 | 2.988 | 15.760 | 1.105 | 1.266 | 1.737 | 0.665 | 0.709 | 1.075 | 0.948 | 0.959 |
| 1957 | 25.397 | 8.719 | 1.496 | 0.559 | 2.498 | 2.058 | 3.718 | 20.450 | 1.362 | 1.608 | 2.214 | 0.867 | 0.955 | 1.410 | 1.244 | 0.381 | 1.440 |
| 1956 | 29.858 | 6.852 | 2.358 | 3.028 | 2.662 | 4.658 | 26.514 | 1.701 | 2.089 | 2.857 | 1.149 | 1.323 | 1.932 | 1.621 | 0.510 | 0.787 | 1.905 |
| 1955 | 24.971 | 10.303 | 8.247 | 3.193 | 5.710 | 33.009 | 2.099 | 2.631 | 3.616 | 1.427 | 1.745 | 2.564 | 2.102 | 0.656 | 1.005 | 1.124 | 2.629 |
| 1954 | 42.086 | 31.374 | 9.195 | 6.921 | 39.898 | 2.593 | 3.311 | 4.730 | 1.795 | 2.243 | 3.453 | 2.939 | 0.858 | 1.277 | 1.449 | 1.115 | 3.831 |
| 1953 | 86.102 | 30.543 | 17.934 | 47.153 | 3.063 | 3.956 | 5.878 | 2.174 | 2.726 | 4.351 | 3.840 | 1.063 | 1.581 | 1.785 | 1.402 | 3.082 | 4.103 |
| 1952 | 96.644 | 58.460 | 117.910 | 3.601 | 4.661 | 7.479 | 2.673 | 3.387 | 5.465 | 4.869 | 1.321 | 1.930 | 2.190 | 1.744 | 3.782 | 3.938 | 3.488 |
| 1951 | 146.355 | 301.944 | 9.484 | 5.422 | 9.103 | 3.292 | 4.113 | 6.905 | 6.306 | 1.617 | 2.331 | 2.656 | 2.142 | 4.667 | 4.950 | 0.757 | 5.388 |
| 1950 | 750.680 | 26.465 | 14.278 | 10.874 | 4.023 | 4.978 | 8.612 | 8.004 | 1.965 | 2.804 | 3.203 | 2.583 | 5.632 | 6.148 | 0.952 | 2.567 | 6.709 |

Table. NSSH. Modelled historic F-values.

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0.000 | 0.000 | 0.001 | 0.010 | 0.025 | 0.048 | 0.067 | 0.084 | 0.092 | 0.108 | 0.138 | 0.110 | 0.092 | 0.093 | 0.093 | 0.093 | 0.093 |
| 2004 | 0.000 | 0.000 | 0.001 | 0.024 | 0.047 | 0.071 | 0.094 | 0.185 | 0.223 | 0.273 | 0.324 | 0.374 | 0.273 | 0.264 | 0.273 | 0.262 | 1.340 |
| 2003 | 0.000 | 0.000 | 0.002 | 0.026 | 0.031 | 0.065 | 0.171 | 0.199 | 0.163 | 0.295 | 0.322 | 0.282 | 0.212 | 0.317 | 0.429 | 0.252 | 0.118 |
| 2002 | 0.000 | 0.000 | 0.013 | 0.016 | 0.048 | 0.182 | 0.269 | 0.156 | 0.245 | 0.259 | 0.227 | 0.233 | 0.259 | 0.358 | 0.542 | 0.241 | 0.281 |
| 2001 | 0.000 | 0.000 | 0.000 | 0.006 | 0.086 | 0.235 | 0.150 | 0.176 | 0.199 | 0.200 | 0.237 | 0.242 | 0.401 | 0.166 | 1.120 | 0.230 | 0.261 |
| 2000 | 0.000 | 0.000 | 0.001 | 0.037 | 0.213 | 0.102 | 0.149 | 0.192 | 0.216 | 0.318 | 0.434 | 0.620 | 0.476 | 1.590 | 8.330 | 0.219 | 0.687 |
| 1999 | 0.000 | 0.000 | 0.001 | 0.040 | 0.082 | 0.135 | 0.148 | 0.188 | 0.232 | 0.342 | 0.463 | 0.256 | 0.793 | 0.236 | 1.430 | 0.208 | 0.085 |
| 1998 | 0.000 | 0.000 | 0.014 | 0.125 | 0.179 | 0.097 | 0.150 | 0.177 | 0.282 | 0.330 | 0.450 | 0.261 | 0.084 | 0.606 | 6.820 | 0.198 | 0.000 |
| 1997 | 0.000 | 0.000 | 0.021 | 0.073 | 0.057 | 0.115 | 0.199 | 0.352 | 0.475 | 0.369 | 0.145 | 0.497 | 0.273 | 1.370 | 0.424 | 0.262 | 0.000 |
| 1996 | 0.000 | 0.000 | 0.010 | 0.006 | 0.037 | 0.115 | 0.274 | 0.340 | 0.373 | 0.033 | 0.074 | 0.140 | 0.463 | 0.514 | 0.001 | 0.303 | 0.000 |
| 1995 | 0.000 | 0.000 | 0.000 | 0.002 | 0.021 | 0.128 | 0.329 | 0.469 | 0.072 | 0.124 | 0.112 | 0.925 | 0.320 | 1.200 | 4.050 | 0.100 | 0.014 |
| 1994 | 0.000 | 0.000 | 0.000 | 0.002 | 0.018 | 0.129 | 0.207 | 0.059 | 0.050 | 0.048 | 0.200 | 0.154 | 0.349 | 0.939 | 2.220 | 0.100 | 0.000 |
| 1993 | 0.000 | 0.000 | 0.000 | 0.004 | 0.030 | 0.082 | 0.027 | 0.019 | 0.031 | 0.078 | 0.075 | 0.000 | 0.001 | 0.008 | 0.000 | 0.059 | 0.000 |
| 1992 | 0.000 | 0.000 | 0.000 | 0.003 | 0.025 | 0.013 | 0.005 | 0.011 | 0.020 | 0.034 | 0.175 | 0.486 | 0.963 | 0.041 | 0.048 | 0.029 | 0.000 |
| 1991 | 0.000 | 0.000 | 0.000 | 0.005 | 0.006 | 0.005 | 0.011 | 0.026 | 0.027 | 0.130 | 0.214 | 0.172 | 0.019 | 1.650 | 1.990 | 0.024 | 0.004 |
| 1990 | 0.000 | 0.000 | 0.006 | 0.035 | 0.008 | 0.008 | 0.026 | 0.024 | 0.053 | 0.436 | 1.430 | 0.056 | 1.680 | 0.413 | 0.022 | 0.153 | 0.266 |
| 1989 | 0.000 | 0.000 | 0.027 | 0.008 | 0.002 | 0.012 | 0.029 | 0.113 | 0.149 | 0.198 | 0.062 | 0.912 | 0.744 | 0.010 | 0.171 | 0.090 | 0.000 |
| 1988 | 0.001 | 0.001 | 0.015 | 0.030 | 0.043 | 0.040 | 0.223 | 0.436 | 0.857 | 0.206 | 1.460 | 1.060 | 0.088 | 0.875 | 4.760 | 0.381 | 0.000 |
| 1987 | 0.002 | 0.004 | 0.010 | 0.028 | 0.030 | 0.293 | 0.254 | 0.461 | 0.268 | 0.639 | 1.060 | 0.093 | 0.903 | 1.520 | 1.520 | 0.417 | 0.000 |
| 1986 | 0.004 | 0.000 | 0.003 | 0.027 | 0.187 | 0.588 | 0.530 | 0.565 | 1.260 | 1.250 | 0.835 | 0.820 | 2.060 | 2.160 | 0.001 | 1.400 | 0.594 |
| 1985 | 0.001 | 0.004 | 0.006 | 0.165 | 0.343 | 0.320 | 0.378 | 0.396 | 0.590 | 0.336 | 0.221 | 0.302 | 0.389 | 0.001 | 0.506 | 0.379 | 0.087 |
| 1984 | 0.005 | 0.000 | 0.010 | 0.070 | 0.074 | 0.121 | 0.083 | 0.085 | 0.059 | 0.116 | 0.112 | 0.035 | 0.000 | 0.294 | 0.242 | 0.070 | 0.000 |
| 1983 | 0.001 | 0.008 | 0.015 | 0.036 | 0.033 | 0.035 | 0.034 | 0.021 | 0.017 | 0.025 | 0.033 | 0.058 | 2.280 | 3.310 | 0.030 | 0.029 | 0.000 |
| 1982 | 0.015 | 0.004 | 0.001 | 0.018 | 0.025 | 0.021 | 0.017 | 0.022 | 0.023 | 0.023 | 0.040 | 0.539 | 0.194 | 0.004 | 0.024 | 0.022 | 0.000 |
| 1981 | 0.012 | 0.003 | 0.009 | 0.011 | 0.018 | 0.019 | 0.020 | 0.018 | 0.026 | 0.092 | 0.692 | 0.384 | 0.024 | 0.021 | 0.019 | 0.028 | 0.000 |
| 1980 | 0.007 | 0.000 | 0.001 | 0.021 | 0.011 | 0.018 | 0.027 | 0.042 | 0.093 | 0.033 | 0.002 | 0.055 | 0.017 | 0.016 | 0.023 | 0.058 | 0.000 |
| 1979 | 0.004 | 0.002 | 0.004 | 0.010 | 0.012 | 0.019 | 0.025 | 0.055 | 0.003 | 0.002 | 0.043 | 0.015 | 0.013 | 0.020 | 0.048 | 0.042 | 0.000 |
| 1978 | 0.005 | 0.002 | 0.001 | 0.017 | 0.028 | 0.037 | 0.115 | 0.003 | 0.743 | 0.069 | 0.012 | 0.011 | 0.017 | 0.039 | 0.035 | 0.079 | 0.000 |
| 1977 | 0.013 | 0.002 | 0.010 | 0.043 | 0.036 | 0.036 | 0.003 | 0.266 | 0.116 | 0.011 | 0.010 | 0.014 | 0.033 | 0.029 | 0.064 | 0.116 | 0.000 |
| 1976 | 0.003 | 0.003 | 0.001 | 0.030 | 0.397 | 0.002 | 0.000 | 0.109 | 0.009 | 0.008 | 0.012 | 0.027 | 0.024 | 0.052 | 0.090 | 0.015 | 0.000 |
| 1975 | 0.016 | 0.002 | 0.001 | 0.157 | 0.222 | 0.319 | 0.189 | 0.038 | 0.014 | 0.010 | 0.023 | 0.021 | 0.043 | 0.072 | 0.012 | 0.019 | 0.000 |
| 1974 | 0.012 | 0.002 | 0.097 | 0.122 | 0.060 | 0.112 | 0.973 | 0.779 | 0.009 | 0.019 | 0.017 | 0.035 | 0.058 | 0.011 | 0.016 | 0.017 | 0.000 |
| 1973 | 0.004 | 0.033 | 0.769 | 0.405 | 0.090 | 0.857 | 1.520 | 1.120 | 1.370 | 0.015 | 0.029 | 0.047 | 0.009 | 0.014 | 1.260 | 0.603 | 0.000 |
| 1972 | 0.842 | 1.830 | 0.987 | 0.098 | 1.290 | 0.956 | 1.820 | 1.250 | 2.980 | 1.780 | 0.039 | 1.570 | 2.070 | 1.110 | 0.332 | 2.110 | 0.000 |
| 1971 | 0.226 | 0.429 | 0.125 | 0.283 | 0.140 | 0.308 | 0.277 | 0.476 | 1.630 | 2.550 | 1.960 | 2.280 | 1.840 | 0.217 | 2.800 | 2.030 | 0.000 |
| 1970 | 0.333 | 0.261 | 1.300 | 0.529 | 1.490 | 0.298 | 0.678 | 1.100 | 1.160 | 1.180 | 1.370 | 1.550 | 3.940 | 1.620 | 0.823 | 1.410 | 0.024 |
| 1969 | 0.094 | 1.710 | 1.870 | 2.050 | 0.259 | 0.756 | 0.599 | 0.347 | 0.419 | 0.600 | 0.626 | 0.612 | 0.867 | 0.394 | 0.843 | 0.578 | 0.015 |
| 1968 | 0.775 | 0.722 | 0.724 | 3.230 | 4.590 | 4.750 | 1.810 | 1.210 | 0.936 | 1.580 | 1.290 | 1.320 | 0.771 | 1.960 | 1.250 | 1.360 | 0.067 |
| 1967 | 0.186 | 1.850 | 0.339 | 0.489 | 1.100 | 0.853 | 1.310 | 1.690 | 1.480 | 1.330 | 1.450 | 1.080 | 1.260 | 1.130 | 1.990 | 1.480 | 0.208 |
| 1966 | 0.128 | 0.742 | 0.244 | 0.308 | 0.405 | 0.529 | 0.593 | 0.749 | 1.310 | 0.976 | 1.340 | 1.060 | 0.543 | 1.370 | 1.480 | 1.190 | 0.280 |
| 1965 | 0.554 | 0.181 | 0.198 | 0.672 | 0.177 | 0.156 | 0.297 | 0.280 | 0.444 | 0.239 | 0.344 | 0.838 | 0.659 | 0.691 | 0.627 | 0.620 | 0.948 |
| 1964 | 0.062 | 0.067 | 0.536 | 0.061 | 0.083 | 0.187 | 0.138 | 0.031 | 0.072 | 0.283 | 0.306 | 0.333 | 0.284 | 0.366 | 0.365 | 0.320 | 0.849 |


| 1963 | 0.046 | 0.964 | 0.483 | 0.123 | 0.058 | 0.042 | 0.031 | 0.069 | 0.145 | 0.070 | 0.243 | 0.214 | 0.263 | 0.274 | 0.344 | 0.257 | 0.156 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1962 | 0.364 | 0.271 | 0.090 | 0.102 | 0.052 | 0.044 | 0.109 | 0.121 | 0.071 | 0.097 | 0.090 | 0.125 | 0.167 | 0.130 | 0.163 | 0.147 | 0.143 |
| 1961 | 0.137 | 0.430 | 0.086 | 0.157 | 0.092 | 0.049 | 0.070 | 0.092 | 0.080 | 0.067 | 0.112 | 0.112 | 0.112 | 0.101 | 0.107 | 0.106 | 0.040 |
| 1960 | 0.108 | 0.147 | 0.733 | 0.759 | 0.168 | 0.104 | 0.091 | 0.098 | 0.080 | 0.128 | 0.136 | 0.145 | 0.214 | 0.193 | 0.204 | 0.139 | 0.084 |
| 1959 | 0.071 | 0.102 | 0.611 | 0.105 | 0.078 | 0.077 | 0.115 | 0.099 | 0.115 | 0.099 | 0.114 | 0.148 | 0.178 | 0.221 | 0.168 | 0.113 | 0.101 |
| 1958 | 1.070 | 0.954 | 1.280 | 0.041 | 0.043 | 0.069 | 0.061 | 0.073 | 0.069 | 0.071 | 0.111 | 0.133 | 0.174 | 0.125 | 0.074 | 0.082 | 0.192 |
| 1957 | 0.369 | 0.896 | 0.262 | 0.046 | 0.176 | 0.084 | 0.069 | 0.111 | 0.059 | 0.089 | 0.093 | 0.116 | 0.147 | 0.121 | 0.121 | 0.100 | 0.151 |
| 1956 | 0.331 | 0.622 | 0.540 | 0.042 | 0.107 | 0.076 | 0.110 | 0.072 | 0.111 | 0.105 | 0.131 | 0.176 | 0.165 | 0.115 | 0.143 | 0.130 | 0.219 |
| 1955 | 0.393 | 0.575 | 0.102 | 0.032 | 0.054 | 0.069 | 0.061 | 0.081 | 0.085 | 0.067 | 0.127 | 0.133 | 0.110 | 0.101 | 0.095 | 0.104 | 0.175 |
| 1954 | 0.507 | 0.436 | 0.158 | 0.042 | 0.040 | 0.061 | 0.080 | 0.118 | 0.080 | 0.101 | 0.148 | 0.185 | 0.118 | 0.089 | 0.104 | 0.130 | 0.132 |
| 1953 | 0.110 | 0.301 | 0.052 | 0.017 | 0.017 | 0.028 | 0.067 | 0.041 | 0.045 | 0.081 | 0.117 | 0.065 | 0.064 | 0.059 | 0.079 | 0.078 | 0.085 |
| 1952 | 0.252 | 0.282 | 0.017 | 0.012 | 0.014 | 0.091 | 0.057 | 0.067 | 0.078 | 0.087 | 0.067 | 0.049 | 0.055 | 0.068 | 0.055 | 0.073 | 0.096 |
| 1951 | 0.018 | 0.040 | 0.068 | 0.001 | 0.047 | 0.058 | 0.044 | 0.084 | 0.109 | 0.053 | 0.039 | 0.043 | 0.056 | 0.060 | 0.079 | 0.070 | 0.050 |
| 1950 | 0.011 | 0.126 | 0.068 | 0.028 | 0.051 | 0.041 | 0.071 | 0.088 | 0.045 | 0.035 | 0.038 | 0.037 | 0.038 | 0.067 | 0.078 | 0.046 | 0.033 |



Figure 3.7.2.1.12. NSSH. Retrospective plot for the assessment in 2004.

Table 3.7.4.2 NSSH. Stock summary table. Biomass in million tonnes.

|  | Recruits | Total | Spawning stock | Fbar |
| :---: | :---: | :---: | :---: | :---: |
|  | Age 0 | biomass | biomass | 5-14 |
| 1950 | 750.680 | 20.013 | 14.359 | 0.058 |
| 1951 | 146.355 | 19.274 | 12.635 | 0.070 |
| 1952 | 96.644 | 20.182 | 11.042 | 0.073 |
| 1953 | 86.102 | 17.419 | 9.457 | 0.066 |
| 1954 | 42.086 | 18.565 | 8.703 | 0.113 |
| 1955 | 24.971 | 15.725 | 9.324 | 0.078 |
| 1956 | 29.858 | 13.799 | 10.934 | 0.110 |
| 1957 | 25.397 | 11.088 | 9.661 | 0.103 |
| 1958 | 23.094 | 9.549 | 8.731 | 0.079 |
| 1959 | 412.478 | 8.076 | 7.200 | 0.113 |
| 1960 | 197.514 | 7.634 | 5.853 | 0.136 |
| 1961 | 76.103 | 7.796 | 4.403 | 0.104 |
| 1962 | 19.003 | 6.765 | 3.443 | 0.146 |
| 1963 | 168.931 | 6.913 | 2.641 | 0.253 |
| 1964 | 93.903 | 6.446 | 2.479 | 0.226 |
| 1965 | 8.491 | 5.935 | 2.996 | 0.278 |
| 1966 | 51.409 | 4.392 | 2.658 | 0.696 |
| 1967 | 3.947 | 3.018 | 1.304 | 1.519 |
| 1968 | 5.187 | 0.982 | 0.318 | 3.493 |
| 1969 | 9.785 | 0.190 | 0.142 | 0.590 |
| 1970 | 0.661 | 0.116 | 0.069 | 1.320 |
| 1971 | 0.236 | 0.130 | 0.032 | 1.525 |
| 1972 | 0.957 | 0.085 | 0.016 | 1.497 |
| 1973 | 12.884 | 0.112 | 0.086 | 1.173 |
| 1974 | 8.631 | 0.160 | 0.091 | 0.114 |
| 1975 | 2.971 | 0.302 | 0.079 | 0.190 |
| 1976 | 10.068 | 0.362 | 0.139 | 0.106 |
| 1977 | 5.095 | 0.429 | 0.288 | 0.111 |
| 1978 | 6.201 | 0.579 | 0.360 | 0.043 |
| 1979 | 12.498 | 0.635 | 0.391 | 0.024 |
| 1980 | 1.474 | 0.748 | 0.475 | 0.034 |
| 1981 | 1.100 | 0.796 | 0.509 | 0.022 |
| 1982 | 2.343 | 0.729 | 0.507 | 0.020 |
| 1983 | 322.362 | 1.086 | 0.579 | 0.029 |
| 1984 | 11.528 | 2.009 | 0.603 | 0.090 |
| 1985 | 35.051 | 5.166 | 0.502 | 0.379 |
| 1986 | 6.041 | 1.815 | 0.401 | 1.074 |
| 1987 | 8.945 | 3.117 | 0.877 | 0.404 |
| 1988 | 25.009 | 3.485 | 2.738 | 0.045 |
| 1989 | 67.357 | 4.067 | 3.335 | 0.029 |
| 1990 | 114.598 | 4.550 | 3.490 | 0.022 |
| 1991 | 318.995 | 5.187 | 3.628 | 0.025 |
| 1992 | 371.421 | 6.208 | 3.496 | 0.029 |
| 1993 | 92.042 | 7.270 | 3.352 | 0.068 |
| 1994 | 29.993 | 8.275 | 3.775 | 0.139 |
| 1995 | 9.773 | 9.061 | 4.592 | 0.240 |
| 1996 | 57.485 | 9.123 | 6.113 | 0.197 |
| 1997 | 37.573 | 9.012 | 7.308 | 0.187 |
| 1998 | 258.857 | 7.841 | 6.564 | 0.168 |
| 1999 | 196.635 | 8.863 | 5.930 | 0.212 |
| 2000 | 45.323 | 7.986 | 4.635 | 0.262 |
| 2001 | 16.362 | 6.798 | 3.878 | 0.213 |
| 2002 | 288.899 | 7.220 | 3.918 | 0.232 |
| 2003 | 160.617 | 8.542 | 5.107 | 0.132 |
| 2004 | 401.287 | 10.180 | 6.513 | 0.119 |

Table 3.9.1.1 NSSH. Input to the short term prediction.

Landings in 2005
Fbar age range:
Total biomass in 2005
Spawning stock biomass in 2005
Fbar in 2005
Part of year before spawning: 0.1

1. million tonnes

5-14, Fbar is weighted with population numbers January 1
11.1161 million tonnes
6.13347 million tonnes 0.1846

| Numbers | Weight | Weight | Weight | Weight | Fraction | Fraction | Fraction |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (billion) | stock | stock | stock | catch | mature | mature | mature | rn |
| 2005 | 2005 | 2006 | 2007 | (kg) | 2005 | 2006 | 2007 |  |
| 0.000 | 0.001 | 0.001 | 0.001 | 0.007 | 0.000 | 0.000 | 0.000 | 0.000 |
| 163.151 | 0.010 | 0.010 | 0.010 | 0.065 | 0.000 | 0.000 | 0.000 | 0.000 |
| 26.549 | 0.047 | 0.050 | 0.050 | 0.143 | 0.000 | 0.000 | 0.000 | 0.002 |
| 19.391 | 0.112 | 0.110 | 0.110 | 0.183 | 0.150 | 0.100 | 0.100 | 0.013 |
| 0.921 | 0.154 | 0.160 | 0.160 | 0.211 | 0.450 | 0.600 | 0.450 | 0.047 |
| 2.070 | 0.233 | 0.249 | 0.246 | 0.236 | 0.900 | 0.900 | 0.900 | 0.106 |
| 7.492 | 0.265 | 0.282 | 0.279 | 0.263 | 1.000 | 1.000 | 1.000 | 0.174 |
| 7.709 | 0.289 | 0.310 | 0.306 | 0.297 | 1.000 | 1.000 | 1.000 | 0.215 |
| 0.614 | 0.317 | 0.339 | 0.335 | 0.342 | 1.000 | 1.000 | 1.000 | 0.239 |
| 0.476 | 0.354 | 0.367 | 0.362 | 0.358 | 1.000 | 1.000 | 1.000 | 0.262 |
| 0.078 | 0.362 | 0.379 | 0.375 | 0.361 | 1.000 | 1.000 | 1.000 | 0.254 |
| 0.124 | 0.411 | 0.416 | 0.411 | 0.382 | 1.000 | 1.000 | 1.000 | 0.288 |
| 0.306 | 0.394 | 0.406 | 0.401 | 0.389 | 1.000 | 1.000 | 1.000 | 0.294 |
| 1.027 | 0.398 | 0.409 | 0.405 | 0.399 | 1.000 | 1.000 | 1.000 | 0.465 |
| 0.551 | 0.413 | 0.423 | 0.418 | 0.420 | 1.000 | 1.000 | 1.000 | 0.666 |
| 0.068 | 0.442 | 0.470 | 0.464 | 0.439 | 1.000 | 1.000 | 1.000 | 0.201 |
| 0.020 | 0.445 | 0.470 | 0.464 | 0.439 | 1.000 | 1.000 | 1.000 | 0.282 |

Table 3.9.1.2 NSSH. Short term prediction and two optional runs.
Short term prediction. Averaging period 10 Averaged

| SSB | Fbar | Landings |  |  |
| :---: | :---: | :---: | :---: | :---: |
| 2006 |  | Biomass $3+$ | SSB |  |
| 6.498 | 0.000 | 0.000 | 2007 | 2007 |
| 6.437 | 0.115 | 0.588 | 11.467 | 8.500 |
| 6.435 | 0.120 | 0.613 | 11.848 | 7.866 |
| 6.432 | 0.125 | 0.636 | 11.824 | 7.840 |
| 6.430 | 0.130 | 0.660 | 11.800 | 7.789 |
| 6.427 | 0.135 | 0.684 | 11.776 | 7.764 |
| 6.424 | 0.140 | 0.708 | 11.753 | 7.739 |
| 6.422 | 0.145 | 0.731 | 11.729 | 7.714 |
| 6.419 | 0.150 | 0.755 | 11.705 | 7.689 |
| 6.417 | 0.155 | 0.778 | 11.682 | 7.664 |
| 6.414 | 0.160 | 0.801 | 11.659 | 7.640 |

Optional 1. Short term prediction Averaging period 10 Averaged - Increase 1 year on 4 year olds

| SSB | Fbar | Landings | Biomass $3+$ | SSB |
| :---: | :---: | :---: | :---: | :---: |
| 2006 |  | 2006 | 2007 | 2007 |
| 6.489 | 0.000 | 0.000 | 12.457 | 8.490 |
| 6.424 | 0.115 | 0.676 | 11.767 | 7.771 |
| 6.421 | 0.120 | 0.704 | 11.739 | 7.741 |
| 6.418 | 0.125 | 0.732 | 11.711 | 7.712 |
| 6.415 | 0.130 | 0.759 | 11.683 | 7.683 |
| 6.412 | 0.135 | 0.786 | 11.655 | 7.654 |
| 6.410 | 0.140 | 0.814 | 11.628 | 7.626 |
| 6.407 | 0.145 | 0.841 | 11.600 | 7.597 |
| 6.404 | 0.150 | 0.868 | 11.573 | 7.569 |
| 6.401 | 0.155 | 0.895 | 11.545 | 7.541 |
| 6.398 | 0.160 | 0.921 | 11.518 | 7.513 |

Optional 2. Catches in 2006 and SSB in 2007 if the fishing mortality is applied to spawning stock biomass

|  | Catch <br> 2006 | SSB |
| :--- | :--- | ---: |
| 2007 |  |  |
| 0.000 | 0.000 | 7.380 |
| 0.115 | 0.758 | 7.296 |
| 0.120 | 0.791 | 7.292 |
| 0.125 | 0.823 | 7.288 |
| 0.130 | 0.856 | 7.285 |
| 0.135 | 0.889 | 7.281 |
| 0.140 | 0.922 | 7.277 |
| 0.145 | 0.955 | 7.274 |
| 0.150 | 0.988 | 7.270 |
| 0.155 | 1.021 | 7.267 |
| 0.160 | 1.054 | 7.263 |

## Blue Whiting.

Table. Blue whiting. Stock numbers at age derived by AMCI.

| Age | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |
| 1 | 3,634,926 | 4,074,635 | 10,886,489 | 18,346,768 | 12,241,432 | 9,473,498 | 8,982,704 | 7,684,033 |
| 2 | 3,994,375 | 2,763,744 | 3,179,573 | 7,637,886 | 12,598,817 | 8,534,761 | 6,686,965 | 6,520,681 |
| 3 | 4,787,469 | 2,956,782 | 2,090,780 | 2,345,821 | 5,324,109 | 8,596,914 | 5,530,846 | 4,489,524 |
| 4 | 3,032,825 | 3,246,249 | 2,105,565 | 1,455,627 | 1,547,086 | 3,343,590 | 4,952,787 | 3,357,315 |
| 5 | 2,421,173 | 1,978,893 | 2,236,277 | 1,414,218 | 921,378 | 952,453 | 1,798,949 | 2,745,985 |
| 6 | 2,309,452 | 1,509,387 | 1,327,705 | 1,461,018 | 863,349 | 546,013 | 496,258 | 968,288 |
| 7 | 1,878,814 | 1,228,032 | 903,621 | 759,007 | 748,554 | 417,017 | 222,861 | 222,973 |
| 8 | 1,799,023 | 999,044 | 735,182 | 516,571 | 388,878 | 361,568 | 170,210 | 100,133 |
| 9 | 1,487,675 | 956,615 | 598,095 | 420,281 | 264,666 | 187,837 | 147,578 | 76,477 |
| 10+ | 3,149,132 | 2,465,584 | 2,048,759 | 1,513,122 | 990,581 | 606,312 | 324,140 | 211,947 |
| Age | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 0 |  |  |  |  |  |  |  |  |
| 1 | 8,938,636 | 22,336,700 | 9,008,614 | 6,160,529 | 5,525,904 | 6,505,224 | 7,617,300 | 20,406,428 |
| 2 | 5,786,361 | 6,601,063 | 16,718,297 | 7,119,122 | 4,769,354 | 4,291,179 | 5,060,322 | 5,919,291 |
| 3 | 4,492,794 | 3,896,371 | 4,504,638 | 12,659,255 | 5,362,474 | 3,612,442 | 3,279,957 | 3,817,265 |
| 4 | 2,841,816 | 2,704,454 | 2,391,453 | 3,252,181 | 9,078,501 | 3,872,129 | 2,628,043 | 2,329,541 |
| 5 | 1,958,331 | 1,572,766 | 1,520,596 | 1,665,696 | 2,242,592 | 6,296,292 | 2,722,504 | 1,788,218 |
| 6 | 1,505,873 | 992,705 | 812,496 | 1,010,437 | 1,105,252 | 1,503,193 | 4,251,225 | 1,761,823 |
| 7 | 444,766 | 626,251 | 412,484 | 490,239 | 610,836 | 685,201 | 950,396 | 2,574,480 |
| 8 | 102,419 | 184,966 | 260,217 | 248,883 | 296,363 | 378,688 | 433,219 | 575,546 |
| 9 | 45,994 | 42,593 | 76,856 | 157,008 | 150,456 | 183,730 | 239,426 | 262,351 |
| 10+ | 132,482 | 74,224 | 48,539 | 75,660 | 140,654 | 180,474 | 230,268 | 284,440 |
| Age | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 0 |  |  |  |  |  |  |  |  |
| 1 | 40,137,768 | 30,806,560 | 19,527,742 | 40,592,636 | 64,009,280 | 34,671,620 | 41,366,872 | 41,950,092 |
| 2 | 15,314,992 | 30,492,660 | 23,208,058 | 15,039,314 | 31,336,016 | 48,849,068 | 26,757,696 | 31,224,204 |
| 3 | 4,353,480 | 11,222,027 | 21,317,802 | 16,517,499 | 10,575,065 | 21,754,324 | 34,914,972 | 18,904,828 |
| 4 | 2,589,741 | 2,919,289 | 6,837,649 | 13,164,899 | 9,881,424 | 6,154,365 | 13,411,988 | 21,161,342 |
| 5 | 1,489,306 | 1,639,443 | 1,640,733 | 3,810,114 | 6,916,746 | 4,981,085 | 3,292,873 | 6,970,656 |
| 6 | 1,083,852 | 899,824 | 882,762 | 915,044 | 1,995,481 | 3,480,721 | 2,660,460 | 1,639,702 |
| 7 | 972,908 | 597,449 | 422,277 | 439,253 | 414,796 | 881,843 | 1,649,291 | 1,216,932 |
| 8 | 1,421,670 | 536,294 | 280,376 | 210,121 | 199,116 | 183,306 | 417,849 | 754,409 |
| 9 | 317,826 | 783,664 | 251,677 | 139,512 | 95,249 | 87,993 | 86,857 | 191,130 |
| 10+ | 301,947 | 341,636 | 528,091 | 388,004 | 239,127 | 147,767 | 111,712 | 90,828 |

Table. Blue whiting. Fishing mortality derived by AMCI.

| Age | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |
| 1 | 0.074 | 0.048 | 0.154 | 0.176 | 0.161 | 0.148 | 0.120 | 0.084 |
| 2 | 0.101 | 0.079 | 0.104 | 0.161 | 0.182 | 0.234 | 0.198 | 0.173 |
| 3 | 0.189 | 0.140 | 0.162 | 0.216 | 0.265 | 0.351 | 0.299 | 0.257 |
| 4 | 0.227 | 0.173 | 0.198 | 0.257 | 0.285 | 0.420 | 0.390 | 0.339 |
| 5 | 0.273 | 0.199 | 0.226 | 0.294 | 0.323 | 0.452 | 0.419 | 0.401 |
| 6 | 0.432 | 0.313 | 0.359 | 0.469 | 0.528 | 0.696 | 0.600 | 0.578 |
| 7 | 0.432 | 0.313 | 0.359 | 0.469 | 0.528 | 0.696 | 0.600 | 0.578 |
| 8 | 0.432 | 0.313 | 0.359 | 0.469 | 0.528 | 0.696 | 0.600 | 0.578 |
| 9 | 0.432 | 0.313 | 0.359 | 0.469 | 0.528 | 0.696 | 0.600 | 0.578 |
| 10+ | 0.432 | 0.313 | 0.359 | 0.469 | 0.528 | 0.696 | 0.600 | 0.578 |
| F(3-7) | 0.310 | 0.227 | 0.261 | 0.341 | 0.386 | 0.523 | 0.462 | 0.431 |
| Age | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 0 |  |  |  |  |  |  |  |  |
| 1 | 0.103 | 0.090 | 0.035 | 0.056 | 0.053 | 0.051 | 0.052 | 0.087 |
| 2 | 0.195 | 0.182 | 0.078 | 0.083 | 0.078 | 0.069 | 0.082 | 0.107 |
| 3 | 0.308 | 0.288 | 0.126 | 0.132 | 0.126 | 0.118 | 0.142 | 0.188 |
| 4 | 0.392 | 0.376 | 0.162 | 0.172 | 0.166 | 0.152 | 0.185 | 0.247 |
| 5 | 0.479 | 0.460 | 0.209 | 0.210 | 0.200 | 0.193 | 0.235 | 0.301 |
| 6 | 0.677 | 0.678 | 0.305 | 0.303 | 0.278 | 0.258 | 0.302 | 0.394 |
| 7 | 0.677 | 0.678 | 0.305 | 0.303 | 0.278 | 0.258 | 0.302 | 0.394 |
| 8 | 0.677 | 0.678 | 0.305 | 0.303 | 0.278 | 0.258 | 0.302 | 0.394 |
| 9 | 0.677 | 0.678 | 0.305 | 0.303 | 0.278 | 0.258 | 0.302 | 0.394 |
| 10+ | 0.677 | 0.678 | 0.305 | 0.303 | 0.278 | 0.258 | 0.302 | 0.394 |
| F(3-7) | 0.507 | 0.496 | 0.221 | 0.224 | 0.210 | 0.196 | 0.233 | 0.305 |
| Age | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 0 |  |  |  |  |  |  |  |  |
| 1 | 0.075 | 0.083 | 0.061 | 0.059 | 0.070 | 0.059 | 0.081 | 0.073 |
| 2 | 0.111 | 0.158 | 0.140 | 0.152 | 0.165 | 0.136 | 0.147 | 0.167 |
| 3 | 0.200 | 0.295 | 0.282 | 0.314 | 0.341 | 0.284 | 0.301 | 0.352 |
| 4 | 0.257 | 0.376 | 0.385 | 0.444 | 0.485 | 0.425 | 0.454 | 0.505 |
| 5 | 0.304 | 0.419 | 0.384 | 0.447 | 0.487 | 0.427 | 0.497 | 0.568 |
| 6 | 0.396 | 0.557 | 0.498 | 0.591 | 0.617 | 0.547 | 0.582 | 0.719 |
| 7 | 0.396 | 0.557 | 0.498 | 0.591 | 0.617 | 0.547 | 0.582 | 0.719 |
| 8 | 0.396 | 0.557 | 0.498 | 0.591 | 0.617 | 0.547 | 0.582 | 0.719 |
| 9 | 0.396 | 0.557 | 0.498 | 0.591 | 0.617 | 0.547 | 0.582 | 0.719 |
| 10+ | 0.396 | 0.557 | 0.498 | 0.591 | 0.617 | 0.547 | 0.582 | 0.719 |
| F(3-7) | 0.310 | 0.441 | 0.409 | 0.477 | 0.509 | 0.446 | 0.483 | 0.572 |

Table. Blue whiting. Summary table by AMCI. Run id 20050829194934.741

SUMMARY TABLE

| Year | Recruits <br> age 1 | SSB | F | Catch <br> SOP |
| ---: | ---: | ---: | ---: | ---: |
| 1981 | 3634925 | 2840186 | 0.3102 | 922980 |
| 1982 | 4074635 | 2340060 | 0.2275 | 550643 |
| 1983 | 10886489 | 1903785 | 0.2608 | 553344 |
| 1984 | 18346767 | 1548446 | 0.3409 | 615569 |
| 1985 | 12241432 | 1651896 | 0.3858 | 678214 |
| 1986 | 9473497 | 1855547 | 0.5231 | 847145 |
| 1987 | 8982703 | 1655472 | 0.4617 | 654718 |
| 1988 | 7684032 | 1469257 | 0.4306 | 552264 |
| 1989 | 8938635 | 1412695 | 0.5067 | 630316 |
| 1990 | 22336699 | 1335867 | 0.4962 | 558128 |
| 1991 | 9008614 | 1803269 | 0.2213 | 364008 |
| 1992 | 6160528 | 2437900 | 0.2242 | 474592 |
| 1993 | 5525903 | 2385790 | 0.2096 | 475198 |
| 1994 | 6505223 | 2354832 | 0.1960 | 457696 |
| 1995 | 7617300 | 2180177 | 0.2331 | 505175 |
| 1996 | 20406428 | 2019352 | 0.3047 | 621104 |
| 1997 | 40137767 | 2062619 | 0.3104 | 639680 |
| 1998 | 30806559 | 2827013 | 0.4408 | 1131954 |
| 1999 | 19527741 | 3434783 | 0.4093 | 1261033 |
| 2000 | 40592635 | 3562036 | 0.4773 | 1412449 |
| 2001 | 64009279 | 4005004 | 0.5093 | 1771805 |
| 2002 | 34671620 | 4881275 | 0.4460 | 1556954 |
| 2003 | 41366871 | 5729501 | 0.4834 | 2365319 |
| 2004 | 41950091 | 5113236 | 0.5725 | 2383503 |

Table. Blue Whiting. Input data for short term prediction.
MFDP version 1
Run: AMCI run 2
Time and date: 15:33 29/09/2005
Fbar age range: 3-7
Recruitment, geometric mean of 1996-2003.

2005

| Age | N |  | M |  | Mat |  | PF |  | PM |  | SWt | Sel |  | CWt |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  | 34118628 |  | 0.2 |  | 0.11 |  | 0.25 |  | 0.25 | 0.049 |  | 0.081 | 0.049 |
| 2 |  | 31918788 |  | 0.2 |  | 0.4 |  | 0.25 |  | 0.25 | 0.073 |  | 0.175 | 0.073 |
| 3 |  | 21639784 |  | 0.2 |  | 0.82 |  | 0.25 |  | 0.25 | 0.094 |  | 0.364 | 0.094 |
| 4 |  | 10888237 |  | 0.2 |  | 0.86 |  | 0.25 |  | 0.25 | 0.108 |  | 0.532 | 0.108 |
| 5 |  | 10452690 |  | 0.2 |  | 0.91 |  | 0.25 |  | 0.25 | 0.129 |  | 0.563 | 0.129 |
| 6 |  | 3234738 |  | 0.2 |  | 0.94 |  | 0.25 |  | 0.25 | 0.153 |  | 0.702 | 0.153 |
| 7 |  | 654252.9 |  | 0.2 |  | 1 |  | 0.25 |  | 0.25 | 0.172 |  | 0.702 | 0.172 |
| 8 |  | 485564.7 |  | 0.2 |  | 1 |  | 0.25 |  | 0.25 | 0.198 |  | 0.702 | 0.198 |
| 9 |  | 301014.8 |  | 0.2 |  | 1 |  | 0.25 |  | 0.25 | 0.224 |  | 0.702 | 0.224 |
| 10 |  | 112503.5 |  | 0.2 |  | 1 |  | 0.25 |  | 0.25 | 0.278 |  | 0.702 | 0.278 |


| Age | N |  | M |  | Mat |  | PF |  | PM |  | SWt | Sel | CWt |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  | 34118628 |  | 0.2 |  | 0.11 |  | 0.25 |  | 0.25 | 0.049 |  | 0.081 | 0.049 |
| 2 | . |  |  | 0.2 |  | 0.4 |  | 0.25 |  | 0.25 | 0.073 |  | 0.175 | 0.073 |
| 3 | . |  |  | 0.2 |  | 0.82 |  | 0.25 |  | 0.25 | 0.094 |  | 0.364 | 0.094 |
| 4 | . |  |  | 0.2 |  | 0.86 |  | 0.25 |  | 0.25 | 0.108 |  | 0.532 | 0.108 |
| 5 | . |  |  | 0.2 |  | 0.91 |  | 0.25 |  | 0.25 | 0.129 |  | 0.563 | 0.129 |
| 6 | . |  |  | 0.2 |  | 0.94 |  | 0.25 |  | 0.25 | 0.153 |  | 0.702 | 0.153 |
| 7 | . |  |  | 0.2 |  | 1 |  | 0.25 |  | 0.25 | 0.172 |  | 0.702 | 0.172 |
| 8 | . |  |  | 0.2 |  | 1 |  | 0.25 |  | 0.25 | 0.198 |  | 0.702 | 0.198 |
| 9 | . |  |  | 0.2 |  | 1 |  | 0.25 |  | 0.25 | 0.224 |  | 0.702 | 0.224 |
| 10 | . |  |  | 0.2 |  | 1 |  | 0.25 |  | 0.25 | 0.278 |  | 0.702 | 0.278 |



[^3]Table. Blue Whiting. Short term prediction.


Input units are thousands and kg - output units are tonnes.


[^0]:    * AMCI assessment

[^1]:    ${ }^{15}$ Reported to the EU but not to the ICES WGNPBW. (Landings of 19,467 tonnes)

[^2]:    Note: Age 0 catches should be include next year.

[^3]:    Input units are thousands and kg - output in tonnes

