International Council for the Exploration of the Sea
Conseil International pour
l'Exploration de la Mer

ICES Advisory Committee on Fishery Management ICES CM 2004/ACFM:24

## Report of the

Northern Pelagic and Blue Whiting Fisheries Working Group

27 April - 4 May 2004
ICES, Copenhagen

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### 1.1 Terms of reference

| Code | TOR | Section |
| :--- | :--- | :--- |
| a) | assess the status of and provide catch options for 2005 for the Norwegian spring-spawning <br> herring stock | 3 |
| b) | assess the status of and provide catch options for 2005 for the blue whiting stock; | 6 |
| c) | assess the status of and provide catch options for the 2004-2005 season for the Icelandic <br> summer-spawning herring stocks | 7 |
| d) | assess the status of capelin in Subareas V and XIV and provide catch options for the <br> summer/autumn 2004 and winter 2005 seasons | 4 |
| e) | provide as detailed information as possible on the age/size composition in different segments of <br> the blue whiting fishery | 6.4 .4 |
| f) | provide information on the species compositions in those fisheries that take appreciable <br> amounts of blue whiting, and on the age/size composition by species of these catches [EC <br> request for information on the industrial fisheries] | Not analysed |
| g) | propose measures to reduce exploitation of blue whiting juveniles and evaluate the potential <br> effect on the stock and the fisheries. The evaluation should include, but not be restricted to the <br> effects of introducing a minimum size and closed areas/seasons; | ACFM spring |
| $03,3.12 .5 \mathrm{~b}$ |  |  |
| h) | continue the evaluation of candidates of harvest control rules for blue whiting | Software not <br> ready |
| i) | provide specific information on possible deficiencies in the 2004 assessments including, at <br> least, any major inadequacies in the data on catches, effort or discards; any major inadequacies <br> in research vessel surveys data, and any major difficulties in model formulation, including <br> inadequacies in available software. The consequences of these deficiencies for the assessment <br> of the status of the stocks and for the projection should be clarified | See within <br> eaph |
| j) | comment on this meeting's assessments compared to the last assessment of the same stock, for <br> stocks for which a full or update assessment is presented | 3.3.1 and <br> 6.4 .5 .2 |
| k) | document fully the methods to be applied in subsequent update assessments and list factors that <br> would warrant reconsideration of doing an update, and consider doing a benchmark ahead of <br> schedule, for stocks for which benchmark assessments are done. |  |

### 1.2 List of participants

| Alexander Krysov | Russia |
| :--- | :--- |
| Ane Iriondo | Spain |
| Asta Gudmundsdottir (Chair) | Iceland |
| Dimitri Vasilyev | Russia |
| Dmitry Prozorkevich | Russia |
| Frans van Beek | Netherlands |
| Hjalmar Vilhjalmsson | Iceland |
| Jan Arge Jacobsen | Denmark |
| Jens Chr. Holst | Norway |
| Manuel Meixide | Spain |
| Mikko Heino | Norway |
| Morten Vinther | Denmark |
| Nikolay Timoshenko | Russia |
| Reidar Toresen | Norway |
| Sergei Belikov | Russia |
| Sigurd Tjelmeland | Norway |
| Sveinn Sveinbjörnsson | Iceland |
| Maurice Clarke | Ireland |

For further details see Annex 1.

### 1.3.1 AMCI

The assessment model AMCI (Assessment Model Combining Information from various sources), version 2.1, was described in the Working Group report in 2002. For assessments in 2003 AMCI version 2.2 (May 2002) has been used. This version is essentially an updated version of AMCI 2.1 where some known problems have been solved but without important changes in functionality. An updated manual was available for the Working Group. The Working Group on Methods on Fish Stock Assessments explored and evaluated AMCI 2.2, together with ISVPA and an array of other assessment models in their meeting in early 2003 (ICES 2003/D:03). The report of that Working Group can be consulted for more details on AMCI. For the assessment in 2004 a new version was used, AMCI 2.3a.

### 1.3.2 SeaStar

The assessment program SeaStar is essentially the same model as used during the 2003 meeting for tuning Norwegian spring spawning herring. The model is documented on the web site www.assessment.imr.no, where the user guide and the Mathematica code can be found, as well as supplementary documentation material. Also, a pdf file of the documentation is available from the author and at ICES. Prior to this meeting the possibility of estimating a survey selection pattern along each cohort was implemented. The possibility of a user-defined catchability function also enables the estimation of survey selection.

SeaStar is statistically based. All terms in the likelihood function express the probability for the observation, where the expectation value is given by the modelled stock and the variance depends on parameters that are estimated together with the other tuning parameters. This avoids some of the subjectivity often found in assessments, where specific weights must be given to the various time series of data used for tuning. However, subjective choices must be made at various points in the assessment, for instance which tuning data that are to be used or whether outliers should be excluded. The diagnostics include ordered CDF values, numerical output of the size of likelihoods terms and plots of fit between data and model that may help in deciding which are the better assumptions to use.

### 1.3.2.1 Tuning

SeaStar is a traditional back-calculating tuning model using a VPA based on Pope's approximation. If needed, solving the catch equation is implemented in case the model should be used for a stock with high fishing mortality. The stock is assessed by running the VPA, which is dependent on the F-values in the last year and the F-values for the oldest true age group. Taking the historic stock as the expectation value in underlying distributions for the observed survey data the probability of observing the survey data is calculated. This probability is referred to as the likelihood function. There is provision for selecting different functions to describe the survey distribution. In the present tuning the gamma distribution with a constant CV is used, in accordance with recent practice. Similarly, the probability of observing the tag return data is calculated and included in the likelihood function. It is assumed that the probability of tag returns, which are rare events, follows a Poisson distribution. At the 2000 meeting also a larval observation series was added, where the probability of observation is based on the spawning stock.

The historic stock is assessed by varying the unknown parameters until the maximum of the likelihood function is reached. The parameters that are varied (free parameters, tuning parameters) are:

Catchabilities for the tuning data
Uncertainty parameters for the data
Tagging survival
Terminal F-values

SeaStar provides for basing the likelihood only on the strongest year classes (here referred to as tuned year classes). Also, only the terminal F values for the strongest year classes may be used as tuning variables. The rationale for this is to stabilise the tuning by avoiding bias from large relative errors in the catch in the terminal year of weak year classes, which mediated by the catchabilities would propagate also to the stronger year classes. The terminal F values of the weak year classes are linearly interpolated between the terminal F values that are tuning parameters. The terminal F values of the fish younger than the youngest tuned year class is linearly interpolated to zero at age -1 . The choice of tuned year classes is subjective. Of importance is to avoid the weakest year classes. The Norwegian spring spawning herring has extremely dynamic recruitment and most often the choice is rather obvious.

In SeaStar it is possible to perform the estimation in separate steps, where during one step the parameters estimated in the previous steps may be used, and new parameters can be estimated. This feature is used for Norwegian spring spawning herring in order to first fit the adult part of the stock to the main tuning series in the Norwegian Sea and along the Norwegian coast. Keeping this part fo the stock fixed the young part of the stock is estimated using survey series from the Barents Sea, which are considered more uncertain. The advantage is that the uncertain data from the Barents Sea then will not influence the estimate of the adult stock, that is of the larger importance in the short term. The adult stock at ages 1 and 2 enter into the likelihood terms for the Barents Sea data, though, in order to provide a better basis for the estimation of catchabilities for those data.

The most important output variable is the estimated spawning stock in the assessment year, which is calculated on the basis of number at age, weight at age and maturity at age at January 1 in the assessment year. Number at age is taken from the VPA by calculating forward one year using the catch information in the last year. Maturation at age in the assessment year is assumed equal to the maturation at age in the last year in the VPA. Weight at age in the assessment year is input data. However, it is assumed that the spawning occurs timeBeforeSpawning part of a year into the assessment year and in order to calculate the decrease until spawning time the same F as in the last year of catch is assumed also to apply for the assessment year. However, for the short term projection the WG also will assume that a fixed catch of catchAssessmentYear million tonnes will be taken in the assessment year, which may correspond to a somewhat different F .

### 1.3.2.2 Analysis of assessment uncertainty using bootstrap

The analysis of assessment uncertainty is done using bootstrap. The assessment is run many times, each time new data sets are generated by resampling from the original data set. Catch, survey data, tagging data and larval data may be resampled separately or jointly by appropriate settings which are asked for when the routine Bootstrap is invoked from the main menu, - see the chapter "Running an assessment/Uncertainty analysis by bootstrapping" in the manual for details.

### 1.3.2.2.1 Surveys

The surveys are resampled from the distribution that is assumed when the likelihood function is constructed, based on the unperturbed surveys. This is done by a call to the routine drawSurveys from the routine doOneRun before the likelihood function is evaluated.

### 1.3.2.2.2 Tagging

The number of tags recovered are sampled from the same distribution as assumed when the likelihood function is evaluated, i.e. Poisson. The number of fish screened for tags is assumed normally distributed with a CV specified in the input data list. The uncertainty in the number of screened fish stems from uncertainty regarding the amount of fish screened and uncertainty in the calculation of number at age screened form biological samples taken from the catch. The number of tagged fish released is also assumed uncertain where a normal distribution with a CV specified in the input data list is assumed.

### 1.3.2.2.3 Catch

The catches are considered certain so there is no distribution from which to draw catch data. The best method would be to base the catch data bootstrap on the biological samples used for distributing the catch on age. However, a possibly large source of error in the age distribution of the catch data comes from using biological samples from one space-time domain on catches from another space-time domain. This is necessary because of inadequate biological sampling of the catch from the countries involved in the fishery. The associated error cannot be dealt with however without implementation of the biological samples from all countries and using a time-space model of the fish distribution. This is an important but large project that ideally should be a joint effort of the countries involved.

## catchBootstrapMethod = useFResiduals

If the variable catchBootstrapMethod has the value useFResiduals a polynom of degree catchPolynomDegree is fitted to the the F-values from the baseline assessment for each year. During bootstrapping Fvalues are resampled from the residuals and new catches are calculated.

## catchBootstrapMethod = useTransfer

If catchBootstrapMethod is useTransfer it is assumed that the error is catch stems from misreading the age by one year. For two neighbouring age groups with number at age of stock1 and stock2 (as based on the unperturbed assessment) the catch to transfer is calculated as:

$$
\text { transferred }=\text { marTransferCoefficient }\left(1.0-\frac{\text { Abs (stock1 }- \text { stock2) }}{(\text { stock1 }+ \text { stock } 2)}\right)
$$

where maxTransferCoefficient is a setting.

## Larvae

As for the surveys, the larval data are resampled from the assumed distribution.

In bootstrap runs first a run with the original data is performed. In the first run the setting perform<bootstrap Tag>Bootstrap must be True and the setting draw<bootstrap tag>Bootstrap must be False, where <bootstrap Tag> is Survey, Catch or Tagging. The first run may then be used as a basis for bias correction of the bootstrap.

The bootstrap replicates contain all information that later may be requested by other assessment programs: historic spawning stocks, historic recruitments and when the program is used for tuning Northeast arctic cod even the cod stockdependent part of the predation by cod on capelin.

The bootstrap replicates may be viewed by the top-level routine showBootstrapEntities, which also is used when the standard output is produced.

### 1.3.2.3 Bias in the assessment due to discarding of old scales

As fish grow older the growth zones get closer and become more difficult to read. Consequently, the age of older fish is more uncertain than the age of younger fish. When the age readers experience that the outer growth zones for a particular fish cannot be distinguished from one another, the data for that fish are discarded. This introduces a bias in the age distribution for a given sample. An alternative way of recording the age can help dealing with this problem (Schweder and Tjelmeland, 2003). Using the new method, when the age reader is uncertain about the age a youngest age is recorded instead of an assumed correct age. The age distribution is calculated using the maximum likelihood method. For each fish it is recorded whether the age is certain by the censor category $d=0$ and whether it is a minimum age by the censor category $\mathrm{d}=1$. The age distribution for all fish in a sample is then estimated using a traditional maximum likelihood approach by minimising:

$$
\sum_{i=1}^{N}\left(\left(1-d_{i}\right) \log \left(p_{a_{i}}\right)+d_{1} \log \left(\sum_{a=a_{i}}^{N} p_{a}\right)\right)
$$

with respect to the relative age frequencies $\mathrm{p}_{\mathrm{a}}$, where a is the age.

Implementing this method in the assessment requires re-reading of a large number of historic data. However, this is a large undertaking and in SeaStar a provisional way of treating this problem has been implemented for Norwegian spring spawning herring. Based on a large number of age data obtained both with the traditional method and the new method the conditional probability of discarding a fish after the traditional method as function of age is found. It can then be shown that the bias corrected age distribution can be expressed as (Schweder and Tjelmeland, 2003):

$$
\frac{n-m}{n\left(1-r_{a}\right)} f_{\mathrm{a}}
$$

where a total of $n$ fish have been read, of which mfish have been discarded. $r_{a}$ is the conditional probabiliy of discarding and $\mathrm{f}_{\mathrm{a}}$ is the age distribution obtained. The bias corrected age distribution must be normalized.

This problem becomes to be especially important this year as the strong 1991 and 1992 year-classes grow into the problematic age range. Earlier, we have dealt with the problem that fish have been transferred from the 1983 year-class to the 1985 year-class, probably because of the same effect.

### 1.3.3 ISVPA

The ISVPA is described in last years WGNPBW report (ICES 2002/ACFM:19). Some changes have been made to the model, and these are described in the report from the meeting of the Methods Working Group in January 2003 (ICES 2003/D:03). For the assessment in 2004 ISVPA version 2004.1 was used (working document by D.A.Vasilyev).

### 1.4 Recommendation

The working group recommends that the time of the meeting of the working group should be changed to autumn. The reason for this is that there are acoustic surveys for herring and blue whiting during late spring and summer, and the inclusion of the information from these surveys, particularly of the strength of the incoming year-classes would improve the predictions.

It is recommended to hold a workshop with the aim to study the possibilities of improving the fish stock assessments by including environmental data.

### 1.5 Where TORs are adressed

| Code | TOR | Section |
| :---: | :---: | :---: |
| a) | assess the status of and provide catch options for 2005 for the Norwegian spring-spawning herring stock |  |
| b) | assess the status of and provide catch options for 2005 for the blue whiting stock; |  |
| c) ${ }^{\text {a }}$ | assess the status of and provide catch options for the 2004-2005 season for the Icelandic summerspawning herring stocks |  |
| d) ${ }^{\text {a }}$ | assess the status of capelin in Subareas V and XIV and provide catch options for the summer/autumn 2004 and winter 2005 seasons |  |
| e) p | provide as detailed information as possible on the age/size composition in different segments of the blue whiting fishery | 4.4 |
| f) ${ }^{\text {p }}$ | provide information on the species compositions in those fisheries that take appreciable amounts of blue whiting, and on the age/size composition by species of these catches [EC request for information on the industrial fisheries] | Not |
| g) ${ }^{\text {p }}$ | propose measures to reduce exploitation of blue whiting juveniles and evaluate the potential effect on the stock and the fisheries. The evaluation should include, but not be restricted to the effects of introducing a minimum size and closed areas/seasons; | $\begin{array}{r} \hline \text { ACFM spring } \\ 03,3.12 .5 \mathrm{~b} \end{array}$ |
| h) cond | continue the evaluation of candidates of harvest control rules for blue whiting | Software not ready |
|  | provide specific information on possible deficiencies in the 2004 assessments including, at least, any major inadequacies in the data on catches, effort or discards; any major inadequacies ine research vessel surveys data, and any major difficulties in model formulation, including inadequacies in available software. The consequences of these deficiencies for the assessment of the status of the stocks and for the projection should be clarified | See within ch chapter 3 |
| j) $\quad \mathrm{c}$ | comment on this meeting's assessments compared to the last assessment of the same stock, for stocks for which a full or update assessment is presented | 3.3 .1 and |
| k) ${ }^{\text {k) }}$ | document fully the methods to be applied in subsequent update assessments and list factors that would warrant reconsideration of doing an update, and consider doing a benchmark ahead of schedule, for stocks for which benchmark assessments are done. |  |

### 2.1 Climate considerations in the Barents Sea

### 2.1.1 Temperature

The Barents Sea is a shelf area, receiving inflow of Atlantic water from the west. The inflowing water demonstrates considerable interannual fluctuations in water mass properties, particularly in heat content, which again influence on winter ice conditions. The variability in the physical conditions is monitored in two sections (Fig. 2.1.1.1). FugløyaBear Island is situated where the inflow of Atlantic water takes place; the Vardø-N section represents the central part of the Barents Sea. In both sections there are regular hydrographic observation, and in addition, current measurements have been carried out in the Fugløya- Bear Island section continuously since August 1997.

Figure 2.1.1.2 shows the temperature and salinity anomalies in the Fugløya-Bear Island section in the period from 1977 to January 2004. Temperatures in the Barents Sea were relatively high during most of the 1990s, and with a continuous warm period from 1989-1995. During 1996-1997, the temperature was just below the long-term average before it turned warm again at the end of the decade. Even if the whole decade was warm, it was only the third warmest decade in the $20^{\text {th }}$ century (Ingvaldsen et al., 2002, Ingvaldsen, 2003).

In January 2003 the temperature was just above the long-term average in the whole Barents Sea, but then the temperature increased quickly until March when it was $0.7^{\circ} \mathrm{C}$ above the long-term mean. From April and the rest of the year, the temperature was $0.5^{\circ} \mathrm{C}$ above the long-term average. In January and March 2004 the temperature was still $0.5^{\circ} \mathrm{C}$ above the average.

In June 2003 the temperature in the Fugløya-Bear Island section was about $1^{\circ} \mathrm{C}$ above the long-term average, and the highest observed since the regular observations started in 1977. The temperature decreased slowly towards the longterm average during summer, but was still $0.6^{\circ} \mathrm{C}$ higher in October. Followed by a more rapid decrease, the temperature in January 2004 was exactly on the average.

The monthly sea temperature series from the Russian Kola meridian transect ( $33^{\circ} 30^{\prime} \mathrm{E}, 70^{\circ} 30^{\prime} \mathrm{N}$ to $72^{\circ} 30^{\prime} \mathrm{N}$ ) begins in 1921 (quarterly values back to 1900). The values were calculated by averaging along the transect and from 0 to 200 m water depth vertically (Bochkov, 1982). The Kola section is strategically placed to monitor the variability in the temperature of the eastern part of the Barents Sea dominated by inflowing Atlantic water masses. Values for 2001-2003 and statistics for the period 1921-1999 are shown in Table 2.1.1.1. These values may e.g. be useful for examination of the effects of extremely cold or warm conditions on cod bioenergetics, growth, recruitment and distribution.

In the beginning of the year the temperature was slightly less than the long-term average. The temperature increased rapidly during spring/summer, and in June the temperature reached the overall maximum value with almost $1{ }^{\circ} \mathrm{C}$ above the long-term average. Compared to 2002 the 2003 temperature followed approximately the same development until August. In August-October the temperature was still above the long-term average, but lower than in 2002. At the end of the year the temperature was back at the long-term average, and actually slightly colder in December. Overall the temperature in the Kola section started similar to the Fugløya-Bear Island section, but the warming period started later and reached higher values in the summer. In the autumn/winter the development in the two sections separated, as the Fugløya-Bear Island section remained constant above the long-term average while the Kola section cooled off towards the long-term average at the end of the year.

In the beginning of 2004 the Kola temperature was 4.47 in January and 4.10 in February, which is respectively 0.6 and 0.7 above the long-term average. In addition the first measurements in March indicates the temperature will stay approximately 0.7 above the long-term average for this month.

### 2.1.2 Ice conditions

Figure 2.1.2.1 shows the ice index for the Barents Sea. The variability in the ice coverage is closely linked to the temperature of the inflowing Atlantic water. The ice has a relatively short response time on temperature change (about one year), but usually the sea ice distribution in the eastern Barents Sea respond a bit later than in the western part. 2003 had a negative ice index, which means more ice than average. This was very surprising since the sea temperature was high. There were two reasons for this. Firstly the really ice melt did not start before mid June, which is about one month later that usual. Secondly, the ice melt during summer was extremely low, most likely due to atmospheric forcing. In 2004 the ice coverage is expected to be the same as in 2003, but the ice index will depend on the ice melting in the summer 2004.

The observed current in the section Fugløya-Bjørnøya is predominantly barotropic, and reveals large fluctuations in both current speed and lateral structure (Ingvaldsen et al. 2002). Based on several years of hydrographic observations, and also by current measurement from a 2 -month time series presented by Blindheim (1989), it was believed that the inflow usually take place in a wide core located in the area $72^{\circ} 30^{\prime}-73^{\circ} \mathrm{N}$ with outflow further north. The long-term measurements that started in August 1997 showed a more complicated structure of the current pattern in the area. The inflow of Atlantic water may also be split in several cores. Between the cores there might be a weaker inflow or a return flow. The outflow area may at times be much wider than earlier believed, stretching from $73^{\circ} 30^{\prime} \mathrm{N}$ south to $72^{\circ} \mathrm{N}$. This phenomenon is not only a short time feature; it might be present for a whole month. These patterns are most likely caused by horizontal pressure gradients caused by a change in sea-level between the Barents Sea and the Arctic or the Norwegian Sea either by accumulation of water or by an atmospheric low or high.

There seems to be seasonality in the structure of the current. During winter the frequent passings of atmospheric lows, probably in combination with the weaker stratification, intensify the currents producing a structure with strong lateral velocity-gradients and a distinct, surface-intensified, relatively high-velocity, core of inflow. During the summer, when the winds are weaker and the stratification stronger, the inflowing area is wider, and the horizontal shear and the velocities are lower. In the summer season there is in inflow in the upper 200 m in the deepest part of the Bear Island Trough.

The time series of volume and heat transports reveal fluxes with strong variability on time scales ranging from one to several months (Fig. 2.1.2.2). The monthly mean volume flux is fluctuating between about 5.5 Sv into and 6 Sv out of the Barents Sea, and with a standard deviation of 2 Sv . The strongest fluctuations, especially in the inflow, occur in late winter and early spring, with both maximum and minimum in this period. The recirculation seems to be more stable at a value of something near 1 Sv , but with interruptions of high outflow episodes. High outflows occurred around April in 1998, 1999 and 2001. In 2000 there was strong outflow in January while in 2002 and 2003 strong outflow was observed in August/September. In the first half of 2003 the inflow was high, which may explain the rapid temperature increase between January and March. The intensity of the flow was reduced during spring and summer. Figure 2.1.2.3 show the variability in the inflow as calculated from a wind driven numerical model. Except for January, it is a good fit with the observations. The model results indicate that the variations in the local atmospheric pressure field may be important for the inflow of Atlantic water to the Barents Sea (Ådlandsvik and Loeng, 1991, Ingvaldsen et al., 2002).

### 2.1.3 Predicting Barents Sea temperature

Prediction of forthcoming environmental conditions, or at least some knowledge on the predictability, is most valuable for projecting the survival of fish through the early life stages, as well as weight and maturity at age. The natural first environmental parameter to try to forecast is sea temperature. The rates of a number of growth-related processes are controlled by temperature (Michalsen et al., 1998). In addition, temperature affects almost all species in the ecosystem, making it an important indicator of changes in fish population dynamics (Daan et al., 1994). Furthermore, the "long memory" of the ocean, as compared to the atmosphere, makes it, at least a priori, feasible to realistically predict ocean temperature much further ahead than the typical weather forecast.

Prediction is, however, complicated by the variation in Barents Sea temperature being governed by processes of both external and local origin operating on different time scales. The volume flux and temperature of inflowing Atlantic water masses as well as heat exchange with the atmosphere, possibly linked to atmospheric teleconnections, is important in determining the temperature of the Barents Sea (Ådlandsvik and Loeng, 1991; Loeng et al., 1992). Thus, both slowly moving advective propagation and rapid barotropic responses due to large-scale changes in air pressure must be considered.

A successful statistical forecasting scheme is subject to the availability of long and continuous time series. We therefore base our prediction on the Kola section time series, the longest below-surface sea temperature series in the region. The climatic variations in the Barents Sea depend mainly on the volume of inflowing Atlantic water (Ådlandsvik and Loeng, 1991), which is influenced by both atmospheric processes and density differences in the ocean itself. In addition, local heat exchange with the atmosphere is important.

The major changes in Barents Sea climate take place during the winter months. The variability in the amount of heat flowing in with Atlantic water masses from the south is particularly high during this season. Furthermore, variability in low-pressure passages and cloud cover has an extra strong influence on the winter atmosphere-ocean heat exchange. The difference in temperature between ocean and atmosphere is highest, but highly variable, at this time of year. The air temperature may at times be 30 degrees lower than the SST. Thus, also with regards to the degree of loss of energy to the atmosphere, this season is decisive.

This seasonal difference is reflected in the merit of simple six months forecasts of Kola-section temperature based on linear regression models. Table 2.1.3.1 (from Ottersen et al., 2000) shows that the predictive value for a specific month based on values from six months earlier varies considerably throughout the year. The tendency found was that of persistence across the spring and summer months being higher than for other seasons, allowing for reasonably reliable forecasts from spring until autumn.

The predictions (Tab. 2.1.3.1) indicate that the temperatures in the southern Barents Sea will be close to average from April to June, followed by a warm period from July to September.

### 2.2 Zooplankton

Sampling of zooplankton on a regular basis began in the Barents Sea in 1979, and since 1986 zooplankton abundance has been monitored at annual surveys during joint Norwegian/Russian 0 -group and capelin surveys during AugustOctober. At this time of the year most of the production has taken place and the measured zooplankton biomass can be considered as the overwintering population of zooplankton. In addition, standard sections Bjørnøya-Fugløya and VardøN (since 1991) are covered on average 6 and 4 times a year respectively. From 2003 onwards, the 0 -group and capelin surveys were merged into a single "ecosystem cruise".

These investigations have provided information on zooplankton (e.g. annual and regional variations in abundance, biomass and species composition) to different research groups at IMR. The results are presented in the annul report at IMR and also at ICES Northern Pelagic and Blue Whiting Fisheries and Arctic Fisheries Working Group meetings. Our main aim in the future is to incorporate zooplankton information in the prognosis of growth of capelin and other important fish species.

Plankton samples are obtained by using WP2 and the MOCNESS (Multiple Opening Closing Net and Environmental Sensing System) plankton nets. The sampling depths in the Barents Sea for the WP2 are from bottom to 0 m and 100 to 0 m . At most stations the MOCNESS nets were towed in oblique hauls from 300-200, 200-150, 150-100, 100-50, 50-25, and $25-0 \mathrm{~m}$. The number of nets varied from about 3 to 8 according to the bottom depth.

The zooplankton samples were usually separated into two halves. One half preserved in $4 \%$ formaldehyde was used for species identification. The second half was size fractionated using sieves of 3 mesh sizes; $180 \mu \mathrm{~m}, 1000 \mu \mathrm{~m}$ and $2000 \mu \mathrm{~m}$ for dry weight measurements.

For each MOCNESS and WP2 profiles the biomass ( $\mathrm{mg} \mathrm{m}^{-3}$ and $\mathrm{g} \mathrm{m}^{-2}$ ) and abundance of individuals (nos. $\mathrm{m}^{-3}$, nos. $\mathrm{m}^{-}$ ${ }^{2}$ ) was calculated, using the depth interval and the volume of water filtered.

The crustaceans form the most important group of zooplankton, among which the copepods of the genus Calanus play a key role in the sub-Arctic and Arctic ecosystems. Calanus finmarchicus is the most important contributor to the zooplankton biomass of the Barents Sea (Melle and Skjoldal, 1998). Krill is another group of crustaceans playing a significant role in the pelagic ecosystem as food for both fish and sea mammals. They appear both in large schools and as continuous layers, often staying deep at daytime and ascending at night. In general, amphipods are ranked third in numerical abundance of zooplankton, far exceeded by copepods and krill. Members of the hyperiid amphipod genus Themisto overwhelmingly dominate amphipod fauna, in the central and northern Barents Sea. In the arctic waters, the larger amphipod T. libellula $(60 \mathrm{~mm})$ is a key species.

### 2.3 Trophic interactions

### 2.3.1 Zooplankton and capelin interactions

In the Barents Sea ecosystem, capelin plays a very important role, on one hand as a major predator and on the other hand as a major prey. Capelin is the main predator on zooplankton, feeding mainly on copepods, krill and amphipods. These studies from the Barents Sea have shown that zooplankton biomass in the Barents Sea in the period 1979-2003 has shown several fold variation among years (Figs 2.3.1.1, 2.3.1.2 and 2.3.1.3). Possible reasons for the large variations are the differences in advective transport and predation pressure.

We also observed an inverse relationship between zooplankton and capelin (Fig. 2.3.1.3) indicating low zooplankton abundance when capelin stock size is high and vice versa. The observations of low zooplankton abundance when capelin stock is large is not surprising as capelin is the most important predator on zooplankton in the Barents Sea ecosystem, and probably exploits most of the secondary production during their feeding season. During periods when the capelin stock was at very low levels the predation pressure on zooplankton was minimum, thus allowing
zooplankton individuals to grow into older age groups. This again probably led to high individual growth rate and rapid recovery of the capelin stock. Our study verifies earlier findings (Dalpadado and Skjoldal, 1996, Drobysheva et al., 2003), which had longer data time series.

### 2.3.2 Zooplankton, capelin and cod interactions

Cod (Gadus morhua) is a major predator on the Barents Sea ecosystem. Growth of young Northeast Arctic cod in the Barents Sea has shown strong fluctuations. The mean length of age 3 cod in the Norwegian winter bottom trawl survey has varied between 28 and 42 cm during the period 1984-2002 (ICES 2003a). Correspondingly, the mean weight at age 3 in this survey has varied between 200 and 800 g . Thus, in order to give predictions of cod stock biomass, it is important to predict size at age and not only abundance at age.

Individual growth in fish depends on density dependent factors such as availability of prey. However growth is also dependent on a series of processes (feeding, metabolism, excretion etc.), which are controlled by temperature (Ottersen et al. 2002; Michalsen et al. 1998).

Diet investigations were carried out on 0,1 and 2 year old Northeast Arctic cod sampled in the Barents Sea during 1984-2002 (Dalpadado and Bogstad, 2004). The purpose of this paper was twofold. First, to investigate variations in the diet of age $0-2$ Northeast Arctic cod in the Barents Sea related to prey abundance. Second, to study how variability in growth of these age groups of cod is related to stomach content and food abundance.

Stomach content analyses showed that the 0 and 1 group cod fed mainly on crustaceans with krill and amphipods comprising up to $70 \%$ of their diet (Fig 2.3.2.1 and 2.3.2.2). Krill (Thysanoessa spp. and M. norvegica) and amphipods (Themisto spp.) were mainly found in cod stomachs sampled in the central and close to the Polar Front region in the Barents Sea where these prey organisms are reported to be abundant in summer.

A shift in the main diet from crustaceans to fish was observed from age 1 to age 2 . The diet of 2 -year-old cod mainly comprised capelin (Mallotus villosus) and other fish, and to a lesser degree, krill and amphipods (Fig 2.3.2.3). Shrimp (mainly Pandalus spp.) was also an important prey in both age 1 and 2 cod. A statistically significant positive relationship was obtained between capelin stock size and the amount of capelin in the diet of 2-year-old cod. Results from this study also show that the larger age 2 cod preyed more on capelin in winter and that larger cod ( $>22 \mathrm{~cm}$ ) prefer larger capelin ( $>12 \mathrm{~cm}$ ). During summer capelin migrates to Arctic waters to the north of the Polar Front to feed. Thus, to a large extent, the main distribution area of cod and capelin do not overlap, during summer. This is possibly the reason why we observed the age 2 cod to feed little on capelin during the summer-autumn period. In warm years e.g. 1992, the distribution of capelin is extended to eastern areas. Age 2 cod feed heavily on capelin during winter, in the eastern Barents Sea. Our results show that in years with low capelin abundance, the cod switched to other prey organisms such as shrimps, krill and amphipods. Similarly, the Icelandic cod also switched to other prey when capelin abundance was low (Magnússon and Pálsson 1989).

A positive significant relationship was also obtained between Total Fullness Index (TFI) and the amount of capelin in the diet and between TFI and the growth of 2 -year-old cod indicating that the growth of age 2 cod is to a large extent dependent on the amount of capelin consumed.

### 2.3.3 Predicting fish recruitment

Predictions of the recruitment in fish stocks are essential for future harvesting. Traditionally prediction methods have not included effects of climate variability. In the following some results of multiple linear regression models are presented, where couplings between climate variables and fish stock variables have been used to model the recruitment of North East Arctic cod, Barents Sea capelin and the Norwegian spring spawning herring. The models are novel, and are still under evaluation in search for better fit (Stiansen et al., 2002, Stiansen et al., 2003a, Stiansen, 2003). However, the fit of the model are encouraging, and the models might at present prove useful as background information in stock assessment. In the 2003 capelin assessment the presented capelin model was incorporated into the 1.5 -year projection.

Four models are presented.

- For the 0 -group log index of North East Arctic cod, with two year prognoses
- For the number of recruits (3 year olds) of North East Arctic cod, with two year prognoses
- For the number of recruits (1 year olds) of Barents Sea capelin, with one year prognoses
- For the number of recruits (3 year olds) of Norwegian spring spawning herring with three year prognoses


### 2.3.3.1 0-group index of North East Arctic cod

A model of the log 0-group index (Stiansen et al., 2003b) based on the NAO (North Atlantic Oscillation, see Hurrell et al., 2003 for an autorative overview of the NAO and it's ecological effects) and SSB (spawning stock biomass) 2 years earlier explains $\sim 50 \%$ of the variation in the 0 -group index in the period 1966-2003 (Fig 2.3.4.1.1).

The correlation between the 0 -group log index and the NAO winter index 2 years earlier is high ( 0.72 for the period 1978-2003). This might be explained through food availability. Melle and Holst (2001) have found a high correlation between NAO and the zooplankton biomass in the Norwegian Sea the following year. This might imply that another year later food supply is still good for cod larvae on their drift along the Norwegian coast. The good recruitment of copepods may also use one year to advect into the Barents Sea, giving a two-year time lag. The condition of the spawning stock in the preceding year may also be an important factor.

It would be more appropriate to use the SSB the same year as the 0 -group index. However there is a high autocorrelation in the spawning stock from one year to another. In order to fully use the time lag of two years in the NAO - 0-group relationship we used the SSB two years earlier as well. The autocorrelation for SSB in the period 19462000 was 0.90 and 0.71 for time lags of one and two years, respectively.

The model is:

$$
0 \text { group }_{t}=0.17 \times N A O_{t-2}+1.91 \times 10^{-6} \times S S B_{t-2}+0.36
$$

where 0 -group is the $\log 0$-group index, $N A O$ the Lisboa-Iceland winter index and $S S B$ the spawning stock biomass (tonnes) from the AFWG assessment 2003 (ICES, 2003a). The subscripts denote the time lag in years. Further details can be found in Stiansen et al. (2002).

The prognoses of the 0 -group index show a medium strong increase in 2004 to a value of 1.5 , followed by a lesser increase to a moderately high value of 1.7 in 2005.

### 2.3.3.2 Recruits of North East Arctic cod

A model of the number of three year old recruits of North East Arctic cod (Stiansen et al., 2003a) based on the Kola temperature, the number of 1 year old cods and the capelin maturing biomass explains $\sim 80 \%$ of the variation in the recruitment (fig. 2.3.4.2.1).

The model is:

$$
\operatorname{Re} c 3_{t}=3.0 \times 10^{8} \times T e m p_{t-3}+0.069 \times \operatorname{Re} c 1_{t-2}+7.1 \times 10^{7} \times \log \left(C a p_{t-2}\right)-1.7 \times 10^{9}
$$

where Rec3 is the number of 3 year olds from the AFWG assessment 2003 with cannibalism (ICES, 2003a), Temp the yearly average temperature between 0 and 200 m in the Kola section three years earlier, Recl is the age 1 index of NEA cod from the Norwegian bottom trawl survey in January/February 2 years earlier and Cap is the maturing biomass (tonnes) of capelin from survey estimate of individuals larger than 14 cm 2 years earlier. The subscripts denote the time lag in years. Further details of the model can be found in Stiansen et al. (2003a).

The prognosis shows a stable medium high recruitment in 2004 and 2005 of $680^{*} 10^{6}$ and $750 * 10^{6}$ individuals, respectively.

### 2.3.3.3 Recruits of Barents Sea capelin

A model of the one year old recruits of Barents Sea capelin (Stiansen, 2003) based on the surface (skin) temperature in the Barents Sea, the 0 -group index and the maturing biomass explains $\sim 70 \%$ of the variation in the recruitment (Fig. 2.3.4.3.1). The model has been calculated for the years 1982-2003 (1981-2002 for the dependent variables), with prognoses for 2004.

The model is:

$$
\operatorname{Re} c_{t}=-39.5 \times \text { skin }_{t-1}+0.45 \times \text { 0roup }_{t-1}+0.096 \times \text { matbio }_{t-1}-61
$$

where Rec is the numbers of recruits in $10^{9}$ (data is survey estimates back-calculated to 1 August), skin the skin temperature from the NCEP reanalysed database average from January to March and over the area between $30-45^{\circ} \mathrm{E}$ and $71-75^{\circ} \mathrm{N}$ one year earlier, 0 -group the capelin 0 -group index one year earlier (in August) and matbio the capelin maturing biomass (survey estimates of fish above 14 cm length) one year earlier. The subscripts denote the time lag in years.

The Model have $\mathrm{R}^{2}=0.72$ and P -value $<0.01$, with all individual P -values $<0.03$. The one-year time lag of the dependent variables gives opportunity of a prognosis one year ahead. Further details can be found in Stiansen (2003). Capelin data is taken from Anon. (2003).

The prognoses show a medium recruitment for 2004, with a value of $315 * 10^{9}$ individuals.

### 2.3.3.4 Recruits of Norwegian spring spawning herring

A model for the number of three year old recruits of Norwegian spring spawning herring using the herring 0 -group log index and the NCEP skin temperature describes $\sim 85 \%$ of the variation in the recruitment (Figure 2.3.4.4.1).

The model is:

$$
\operatorname{Re} c_{t}=4.54 \times \text { skin }_{t-3}+16.3 \times \text { group }_{t-3}-24.4
$$

where Rec is the number (in $10^{9}$ ) of 3 year old recruits of Norwegian spring spawning herring from the WGNPBW 2003 SEASTAR assessment (ICES, 2003b), skin the NCEP skin (sea surface) temperature in degree C in the Norwegian Sea ( $64-70^{\circ} \mathrm{N}, 6^{\circ} \mathrm{W}-8^{\circ} \mathrm{E}$ ) averaged from January to March 3 years earlier and 0 group the 0 -group log index of herring larvae from the survey in the autumn 3 years earlier. The subscripts denote the time lag in years. Further details can be found in Stiansen et al. (2002).

The dominant variable in the model is the 0 -group index, which has a correlation coefficient of 0.89 with the Recruitment ( 3 years later). When the model was tested on the 0 -group index alone it gave an $\mathrm{R}^{2}$ of 0.81 . Still the model explained $4 \%$ more of the variability when adding the skin temperature.

The prognosis shows a medium recruitment level for 2004, with a continuous increase for 2005-2006 to a medium high level.

### 2.3.3.5 Key points and summary recruitment

Models, based on climate and fish parameters, for prediction of recruitment have been given for 0 -group index and the number of three year old fish for North East Arctic Cod, for the number of one year old fish for Barents Sea Capelin, and for the number of three year old fish for Norwegian spring spawning herring. Table 2.3.4.5.1 gives an overview of the predicted future recruitment.

- The 0-group index of North East Arctic cod will have a moderately strong increase in 2004, followed by a smaller increase in 2005, giving a medium strong recruitment in both years.
- The number of recruits of North East Arctic cod will be at a medium-high level in 2004-2005.
- The number of recruits of Barents Sea capelin will be at a medium level in 2004.
- The number of recruits of Norwegian spring spawning herring will be at a medium level in 2004 and continuously increase in 2005-2006 to a medium high level.


### 2.4.1 Hydrography and climate

The Nordic Seas during the last decades have been characterized by increased input of Arctic waters. The Arctic waters to the Norwegian Sea are mainly carried by the East Icelandic Current and also to some extent by the Jan Mayen Current. During periods of increased Arctic water input, the western extension of Atlantic water is moved eastward. As a result, over the last 25 years the southern and western Norwegian Sea has become colder and fresher while the eastern Norwegian Sea is warmed. Atmospheric forcing drives this trend. From mid 1960's the winter North Atlantic Oscillation index (NAO) has increased to beginning of the 1990s followed by a reduction to the long-term-mean (Fig. 2.4.1.1). In winter 2004 the index was lower than normal. NAO as it is used here is the normalised air pressure difference at sea level between Lisbon, Portugal and Reykjavik, Iceland and is an indicator of the strength of the westerly winds into the Norwegian Sea. A high NAO index (i.e. stronger westerly winds) will force Atlantic and Arctic waters more eastward.

The Institute of Marine Research, Norway, has measured temperature and salinity in three standard sections in the Norwegian Sea almost regularly since 1978. The sections are 1) the Svinøy section which runs NW from $62.37^{\circ} \mathrm{N}$ at the Norwegian coast, 2) the Gimsøy section which also runs NW from the Lofoten Islands and 3) the Sørkapp section which is a zonal section at $76.33^{\circ} \mathrm{N}$ just south of Svalbard.

Figure 2.4.1.2 shows the development in temperature and salinity in three different sections from south to north in the Norwegian Sea. During the last 7 years the temperature and salinity in the Svinøy section have been above the long-term-mean while they were about average in the Gimsøy and Sørkapp sections. Unfortunately some data are missing in the Gimsøy and Sørkapp sections. In 2003 the salinity in the Svinøy section had the largest value in the time series, about 0.08 above normal. The temperature was the next largest in the time series, about $0.9^{\circ} \mathrm{C}$ above normal. Only in 2002 was the temperature higher.

Figure 2.4.1.3 shows time series of temperature and salinity during the spring in the Svinøy and Gimsøy sections from 1978 to 2003. The values are calculated using the same procedure as mentioned above. The low salinities in 1978 and 1979 are a result of the Great Salinity Anomaly during the 1970 's. In 1994 a large salinity anomaly comparable with the anomaly in 1978 and 1979 was seen in the Svinøy section. The temperature was also a minimum that year. The 1994 anomaly was a result of increased influence of Arctic water from the East Icelandic Current. In 2002 the salinity and temperature increased considerable in the Svinøy section to highest observed value for the time series. The condition in 2003 remained approximately unchanged compared to 2002 . In the Gimsøy section there were instead a reduction in temperature and salinity for 2002 but they increased again in 2003. The salinity was then the highest since 1985.

The area of Atlantic water (defined with $S>35.0$ ) in the Svinøy-section has been calculated. The mean temperature within the limited area has also been calculated, and the results for both spring and summer are shown in Figure 2.4.1.4. There are considerable variations both in the area of Atlantic water distribution and its temperature. The distribution area of Atlantic water has decreased since the beginning of 1980s, while the temperature has shown a steady increase. Since 1978 the Atlantic water has been about $0.6^{\circ} \mathrm{C}$ warmer. During the years 1992-1995 the area was much lower than average for both seasons. In 1997-1999 there was a warm period followed by a substantial drop in temperature in 2000. Then in 2002 the temperature increased considerable and in 2003 it had the largest value in the time series. The temperature was in 2003 about $0.7^{\circ} \mathrm{C}$ higher than the long-term-mean for summer. The area of Atlantic water in 2003 increased also and had the largest value since 1987.

During research cruises in May with the aim of estimating the pelagic stock hydrographic observations are also taken, covering most of the Norwegian Sea. In 2003 there was a larger westerly distribution of Atlantic water than normal and also compared to 2002 (not shown). At 100 m depth the temperature was about $0,5^{\circ} \mathrm{C}$ above normal and in some areas 1,5 ${ }^{\circ} \mathrm{C}$ above normal.

## Conclusions:

- The temperature in the Svinøy section has been the highest ever in 2002 and 2003.
- The winter NAO index in 2004, calculated for December-February, was lower than normal.
- The westerly distribution of Atlantic Water increased in 2003 and a relatively large westerly distribution is also expected in 2004.
- The averaged temperature of the Atlantic Water in the Svinøy section has increased with approximately $0.6^{\circ} \mathrm{C}$ since 1978.


### 2.4.2 Phytoplankton

The development of phytoplankton in the Atlantic water is closely related to the increase of incoming solar irradiance during March and to the development of stratification in the upper mixed layer due to warming. The Institute of Marine Research, Norway, started in 1990 a long-term study of the mechanisms controlling the development of phytoplankton at Ocean Weather Station Mike situated at $66^{\circ} \mathrm{N}, 2^{\circ} \mathrm{E}$.

Figure 2.4.2.1 shows the development of the phytoplankton bloom for 2003 expressed as chlorophyll a concentration at the surface. In previous years there has been a marked difference in the time when the spring bloom reached its maximum. In 1997 the spring bloom reached its maximum 20 May (day of the year 140), in 1998 about one month earlier 18 April (day of the year 108). The timing of the bloom in 1999 was similar to that in 1998, but did not show the same high maximum in chlorophyll. This may be related to the weekly measurements in 1999, as opposed to daily measurements in 1997 and 1998. On the other hand, weekly measurements prior to 1997 have revealed pronounced maxima in chlorophyll. The reason for the low algal biomass in 1999 may have been early and strong grazing from a large over-wintered zooplankton stock. In all these years a strong peak has characterized the bloom. The situation in 2001 was different to previous years. First, the spring bloom started somewhat later (first week of May) compared to 1998 and 1999 and was followed by relatively moderate chlorophyll concentrations culminating with a major peak in the first week of June. Also a distinct early autumn bloom was observed in the middle of August. In 2002 the springbloom started to develop in the middle of April reaching its maximum at the end of April, resulting in one of the earliest bloom second only to the bloom in 1998. The 2003 bloom also maintained relatively high chlorophyll concentrations for about a month after the first peak on May 8 to decrease rapidly afterwards. After the main spring bloom four other smaller blooms were observed throughout the summer and early autumn.

The development of the phytoplankton prior to the spring bloom may be separated into two phases. The first phase, from day 1 to about day 50, is characterised by extremely low phytoplankton biomass expressed as chlorophyll $a$. This is the winter season during which phytoplankton growth is mainly limited by the low incoming irradiance typical of this period. The second phase, from about day 50 to day 100 , is characterised by a gradual increase of phytoplankton biomass but without reaching bloom conditions. This is the pre-bloom phase during which the increase in biomass is related to the increase in incoming irradiance and the lack of a bloom is due to the deep upper mixed layer still present at this time.

Figure 2.4.2.2 shows the extension in time for these two phases and the timing of the spring bloom for the period 19912003. In a "normal" year the winter season extends to about 2 March. The pre-bloom phase extends on average from the 2 March to 16 April. The spring bloom starts normally on 16 April and reaches its maximum on 21 May, but the year-to-year variations are much larger than those of the previous phases. From 1991 to 1995 the trend was towards earlier spring blooms. This trend was broken in 1996, and thereafter year-to-year variability in the timing of the bloom has been greater but with a trend towards earlier blooms again after 2001.

## Conclusions:

- The phytoplankton bloom in 2003 developed earlier than the average since 1991, third only to the 1998 and 2002 blooms.
- Chlorophyll $a$ concentrations first peaked in the first week of May 2002 and were maintained at relatively high levels until the first week of June resulting in the longest bloom in the time series. This could, as in 2002, have been the result of a relax in the grazing pressure.
- During summer and early autumn several peaks of relatively high chlorophyll a concentration were observed indicating a strong variability in minor blooms.


### 2.4.3 Zooplankton

Zooplankton biomass distribution in the Norwegian Sea has been mapped annually in May (since 1995) and in July (since 1994). Zooplankton samples for biomass estimation were collected by vertical net hauls (WP2) or oblique net hauls (MOCNESS). In the present report zooplankton samples from the upper 200 m are analysed. Total zooplankton biomass ( g dry weight $\mathrm{m}^{-2}$ ) in May was averaged over sampling stations within three water masses, Atlantic water (defined by salinity $>35$ at 20 m depths), Arctic water (salinity $<35$, west of $1.4^{\circ} \mathrm{E}$ ) and Coastal water (salinity $<35$, east
of $1.4^{\circ} \mathrm{E}$ ) (Fig. 2.4.3.1). In Atlantic and Arctic water masses zooplankton biomass decreased to a minimum in 1997. Thereafter zooplankton biomass increased again and has remained relatively high except for a temporary reduction in 2001. Due to reduced cruise time the Arctic water mass was not sampled in 2001. For the first time in 2002, the biomass in Atlantic water equalled the biomass in Arctic water. In 2003 zooplankton biomass in Arctic water, again, was much higher than in Atlantic water. However, the confidence limits were wide due to few sampling stations in this water mass. In the Coastal water mass, which includes the Norwegian continental shelf and slope waters influenced by Norwegian coastal water, the trend was different with a general increase towards a maximum in 1998 and a decrease the following years. Biomass increased again in 2002, and reached the highest value during the time series. In 2003 Biomass was lower, but still rather high compared to the biomass in Atlantic water.

In July the total zooplankton biomass ( g dry weight $\mathrm{m}^{-2}$ ) in the upper 200 m was calculated by integrating biomass at sampling stations within a selected area in the central and eastern Norwegian Sea. In 2003 average biomass for the area could not be calculated since the survey area was moved to far to the north. There is no obvious trend in the zooplankton biomass in July since 1994 (Figure 2.4.3.2).

## Conclusions:

- Average zooplankton biomass in Atlantic water masses of the Norwegian Sea in May 2003 was lower than average for the time series.


### 2.4.4 Herring growth and food availability

Individual growth of the Norwegian spring spawning herring, as measured by condition or length specific weight after the summer feeding period in the Norwegian Sea, has been characterised by large fluctuations during the 1990's (Fig 2.4.4.1). During 1991 and 1993 individual condition was good, but from 1994 on the condition of the herring started to decline and by 1997 it reached the lowest level during the 1990's. The level observed in 1997 corresponds with the absolute long-term low level observed during the period 1935-1994 (Dr. scient. thesis J.C. Holst 1996, University of Bergen). Following a recovery during 1998 and 1999, the condition of the herring decreased again. In 2001, 2002 and 2003 the condition remained at a low level.

Since 1995, when the large-scale migration pattern of the herring has been mapped during two annual cruises, May and July-August, the herring have been feeding most heavily in Atlantic water, and the herring condition index obtained after the feeding period in the Norwegian Sea is related to average zooplankton biomass of Atlantic water (Fig. 2.4.4.2). To improve this relationship herring feeding areas should be defined more precisely, because large variations in herring migration routes and in zooplankton distribution have been observed over the years.

Conclusions:

- Herring condition was lower than average for the time series in 2003.
- There is a weak relationship between zooplankton biomass in May and herring condition in the autumn during the years 1995-2003.


### 2.4.5 Predictions for zooplankton biomass and herring feeding conditions

A factor possibly governing zooplankton biomass is the size of the zooplankton spawning stock, or the size of the overwintering population. Zooplankton biomass in July represents the mixed population of zooplankton species at the start of the over-wintering. A linear regression of the biomass in July on the biomass in May the following year explains $\sim 63 \%$ of the total variation (Fig. 2.4.5.1). Average biomass in July 2002 suggested that zooplankton biomass in May 2003 should be close to average as well (Fig. 2.4.5.1). This turned out to be a correct prognosis. However, the time series is short, the variability is large and there is no trend in the July zooplankton biomass that could be related to the trend observed in the May data. Thus, this time series should be expanded before it is used for prediction.

The North Atlantic Oscillation index (NAO), is a proxy for the strength and duration of southwesterly winds, and is correlated with the inflow of Atlantic water to the Norwegian Sea. In the Norwegian Sea the average biomass of zooplankton in Atlantic water in May is strongly correlated with the average NAO for the March-April period the previous year (Fig. 2.4.5.2). March-April is the period when the primary production in the Norwegian Sea is initiated and the major reproductive period for many important zooplankton species such as Calanus finmarchicus and krill. The one-year lag in the relationship may be because we in May mainly measure the size of the previous years overwintering stock, that is the previous years production and the present years spawning stock. The biomass for May 2003 was
predicted at 12.98 g dry weight $\mathrm{m}^{-2}$, based on the NAO for 2002. The observed biomass in 2003 was 10.8 g dry weight $\mathrm{m}^{2}$. Thus, the prognosis underestimated the biomass. Based on the same relationship the biomass for May 2004 is estimated at 10.5 g dry weight $\mathrm{m}^{-2}$.

Biomass (yr2) $=2.67^{*}$ NAO yr $1+11.38$ (1)
$R^{2}=0.75, P=0.003$ The time series for the herring condition index has been calculated for the period from 1991 to 2001. A correlation analysis of herring condition on the two-months average of the NAO indices showed that the relationship was strongest between herring condition and the NAO during the March-April period (Fig. 2.4.5.3). The herring condition index for 2003, based on the NAO for 2002, was predicted to be 0.833 . The observed condition for 2003 was 0.806 , thus lower than predicted. The prediction for 2004 is 0.812 , somewhat higher than in 2003.

Condition (yr2) $=0.023 * N A O$ yrl $1+0.820$ (2)
$R^{2}=0.51, P=0.006$ Conclusions:

- A direct, but weak, relationship between zooplankton biomass in July and the zooplankton biomass in May the following year is suggested by the time series from 1994 to 2002.
- The average NAO for March-April the previous year is directly related to zooplankton biomass in May and herring condition in the autumn.
- The March-April NAO index for 2003 predicts zooplankton biomass at $10.5 \mathrm{~g} \mathrm{~m}^{-2}$ in May 2004 and the herring condition index at 0.806 in the autumn 2004.


### 2.5 Icelandic waters

### 2.5.1 Hydrography and climate

As Iceland is situated at a meeting place of warm and cold currents its waters are characterised by highly variable conditions especially in area north and northeast of the country. Heat and salt content in those waters depend on the strength of Atlantic inflow through the Denmark Strait and the variable flow of polar water from the north with the East Icelandic Current. South and west of Iceland fluctuations are smaller.

Climatic conditions in the North Atlantic improved around 1920 and remained rather warm until the mid-1960s, when they deteriorated. In the area north and east of Iceland temperature and salinity declined sharply in 1965 and these severely cold conditions lasted until 1971. After that, climatic conditions off north and east of Iceland improved, but were variable and years have alternated with cold years (Fig. 2.5.1.1).

Salinity and temperature increased in 1997 west of Iceland and have remained high. This increase in the Atlantic character of the Irminger Current reached into the northern area and peaked in 1999 and prevailed until winter and spring 2002 when a rather short period of polar influence was observed. In summer 2002 a persistent inflow of Atlantic water started and kept on throughout the year 2003 with little winter cooling which resulted in a record year in temperature and salinity north and east of Iceland. Observations in February 2004 showed continued influence of Atlantic water in the northern and eastern area with stronger winter cooling and high salinities.

### 2.5.2 Zooplankton

In the area north of Iceland, zooplankton biomass tends to be higher in years with strong inflow of Atlantic Water than in years when Atlantic inflow is weak, and lower salinity in the surface layers slows or prevents vertical mixing. A strong inflow of Atlantic water to the north Icelandic area was observed both during November 2002 and February/March 2003. The relatively high zooplankton biomass off the central north coast in spring 2003 is in line with this (Fig 2.5.2.1).

In spring 2003, the zooplankton biomass for the whole Icelandic area was slightly below the long-term average. West of Iceland zooplankton biomass was near average, but slightly below the long term mean south and east of Iceland. The copepod C. finmarchicus was generally the dominant zooplankton species in the offshore areas, except in the Arctic East Icelandic current northeast of Iceland, where the arctic copepod C. hyperboreus dominated the biomass.

As mentioned above, a continued strong inflow of Atlantic water to the north Icelandic area was observed both during the November 2003 and February 2004 surveys. This indicates that zooplankton biomass will be above average in north Icelandic warts is spring and summer 2004.

### 2.6 Hydrography of the waters west of the British Isles

Hydrographic data have been collected during surveys in the spawning season of blue whiting in spring. Data from 1983 to 2004 are summarized in Figure 2.6.1. These represent mean values from a box with limits $52^{\circ}$ to $54^{\circ} \mathrm{N}$ and $16^{\circ}$ to $14^{\circ} \mathrm{W}$ (west of the Porcupine Bank) and from depths $50-600 \mathrm{~m}$ wherever bottom depth is 600 m or more. Both salinity and temperature show an increasing trend with the values in 2002-2004 being among the highest ones observed. This trend coincides with an increasing trend in recruitment of blue whiting. However, it is not know whether there is a causal relationship between hydrographic conditions and recruitment of blue whiting.

### 2.7 Present and potential application of ecological considerations in stock assessments.

The efforts put into the preparation of the ecological consideration chapter has been increasing and is at present significant. The actual use of these data in the stock assessment is, however, not proportional to the efforts made in compiling the chapter. This has obviously resulted in frustrations as those preparing the chapter have not prioritised to join the meeting. Presently the only place where these data are taken into use is in the short term prognosis of the herring stock where the condition factor is predicted one year ahead. The use of this method has given better predictions than what would be the case than for instance using the mean of historic values.

There may be many causes for the slow progress in including this type of data in the stock assessments and projections. Inclusions of new data would in many cases call for significant changes in the modelling framework, the persons carrying out the modelling are not the same as those doing the ecologic considerations, the way WGNPBW work may not necessarily be beneficial for facilitating and speeding up such a process and there may not necessarily be obvious and simple ways in including these data in the assessments.

Some ways of including ecologic considerations into the assessments could be mentioned; one example being a revision of the natural mortality of the NSSH, in particular of the younger ages which at present are set to a constant value from age 1 to 3 . Another possibility would be to include the models for predicting recruitment indexes (this chapter) into the predictions.

To speed up the process towards a better integration of the ecological considerations in the assessments it is recommended that a sub group of the WGNPBW meets in Bergen during the autumn of 2004 with the aim of exploring and presenting possible ways for keeping this important process on its tracks and speeding it up.

Table 2.1.1.1. Kola section monthly temperature statistics. Temperature values for 2001, 2002 2003, long-term (19211999) mean, minimum and maximum temperatures and standard deviations are given for each calendar month for the 0200 m depth interval.

| Month | 2003 | 2002 | 2001 | Mean | Minimum | Maximum | Std Dev |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| JANUARY | 3.7 | 3.8 | 4.6 | 3.88 | 2.70 | 5.00 | 0.47 |
| FEBRUARY | 3.3 | 3.6 | 4.2 | 3.44 | 1.80 | 4.70 | 0.51 |
| MARCH | 3.3 | 3.4 | 3.7 | 3.12 | 2.00 | 4.20 | 0.48 |
| APRIL | 3.3 | 3.3 | 3.7 | 2.94 | 1.50 | 3.90 | 0.51 |
| MAY | 3.8 | 3.6 | 3.7 | 3.09 | 1.70 | 4.10 | 0.52 |
| JUNE | 4.5 | 4.2 | 4.3 | 3.56 | 2.30 | 4.50 | 0.51 |
| JULY | 4.8 | 4.9 | 4.7 | 4.18 | 3.00 | 5.20 | 0.52 |
| AUGUST | 4.9 | 5.5 | 5.0 | 4.67 | 3.50 | 5.60 | 0.52 |
| SEPTEMBER | 5.3 | 5.7 | 5.1 | 4.91 | 3.80 | 5.90 | 0.48 |
| OCTOBER | 5.2 | 5.3 | 5.2 | 4.91 | 3.40 | 6.00 | 0.51 |
| NOVEMBER | 4.8 | 5.0 | 4.8 | 4.69 | 3.50 | 5.80 | 0.48 |
| DECEMBER | 4.2 | 4.5 | 4.7 | 4.32 | 3.50 | 5.50 | 0.43 |

Table 2.1.3.1. Linear regression models for monthly $0-200 \mathrm{~m}$ temperature values in the Kola section based on corresponding temperatures from six months earlier. The equations are derived from data from January 1921 to February 1997. All coefficients of determination $\left(\mathrm{R}^{2}\right)$ are significant at the $5 \%$ level.

| Predicted | Prediction | Equation | $\mathrm{R}^{2}$ | Prediction |
| :---: | :---: | :---: | :---: | :---: |
| from month (x) | for month (y) |  |  | 2004 |
| July | January | $\mathrm{y}=1.77+0.50 \mathrm{x}$ | 0.31 |  |
| August | February | $\mathrm{y}=1.05+0.51 \mathrm{x}$ | 0.27 |  |
| September | March | $\mathrm{y}=0.95+0.44 \mathrm{x}$ | 0.21 |  |
| October | April | $\mathrm{y}=0.53+0.49 \mathrm{x}$ | 0.25 | 3.1 |
| November | May | $\mathrm{y}=0.74+0.50 \mathrm{x}$ | 0.22 | 3.1 |
| December | June | $\mathrm{y}=0.98+0.60 \mathrm{x}$ | 0.25 | 3.5 |
| January | July | $\mathrm{y}=1.60+0.67 \mathrm{x}$ | 0.36 | 4.6 |
| February | August | $\mathrm{y}=2.37+0.67 \mathrm{x}$ | 0.41 | 5.1 |
| March | September | $\mathrm{y}=2.67+0.72 \mathrm{x}$ | 0.49 | $5.4^{*}$ |
| April | October | $\mathrm{y}=2.71+0.75 \mathrm{x}$ | 0.55 |  |
| May | November | $\mathrm{y}=2.91+0.57 \mathrm{x}$ | 0.38 |  |
| June | December | $\mathrm{y}=2.71+0.45 \mathrm{x}$ | 0.29 |  |

* Preliminary, based on the assumption that the March 2004 temperature will be 0.7 above the long-term average, as indicated by measurements in the beginning of March.

Table 2.3.4.5.1. The table gives an overview of the different models, with prognoses estimates of the variable in question. The given month indicate when the prognoses can be extended for another year.

| Species | Variable | Prognoses <br> year | Prognoses <br> available | Prognosis <br> $\mathbf{2 0 0 4}$ | Prognosis <br> $\mathbf{2 0 0 5}$ | Prognosis <br> $\mathbf{2 0 0 6}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| North East <br> Arctic cod | 0 -group, log <br> $(0$ year olds) | 2 | November | 1.47 | 1.65 | X |
| North East <br> Arctic cod | Recruits (3 <br> year olds) | 2 | January | $679 * 10^{6}$ | $747^{*} 10^{6}$ | X |
| Barents Sea <br> capelin | Recruits (1 <br> year olds) | 1 | November | $315^{* 10^{9}}$ | X | X |
| Norwegian <br> spring <br> spawning <br> herring | Recruits (3 <br> year olds) | 3 | November | $3.0^{* 10^{9}}$ | $9.0^{*} 10^{9}$ | $12.1^{*} 10^{9}$ |



Figure 2.1.1.1 Standard sections and fixed oceanographic stations worked by Institute of Marine Research, Bergen. The University of Bergen is responsible for station M, while the Kola section is operated by PINRO, Murmansk (Sjøtun, 2004).


Figure 2.1.1.2 Temperature anomalies (upper panel) and salinity anomalies (lower panel) in the section Fugløya - Bear Island (Sjøtun, 2004).


Figure 2.1.2.1. Ice index for the period 1970-2003. Positive values means less ice than average, while negative values show more severe ice conditions (Sjøtun, 2004). Transport and current.

Volume flux into the Barents Sea


Figure 2.1.2.2 Total volume flux across the section Norway-Bear Island All data have been low pass filtered over 30 days (Sjøtun, 2004).


Figure 2.1.2.2 Modelled flux anomalies in 2002 through the section between Norway and Bear Island (Sjøtun, 2004).


Figure 2.3.1.1 Area division for the Barents Sea used in figure 2.3.1.2.


Figure 2.3.1.2. Average zooplankton biomass $\left(\mathrm{g} \mathrm{m}^{-2}\right)$ for the 7 different areas in the Barents Sea during 1988-2003 (Hassel, 2004).


Figure 2.3.1.3 Average zooplankton biomass $\left(\mathrm{g} \mathrm{m}^{-2}\right)$ together with biomass of one year and older capelin (million tonnes) during 1973-2003, in the Barents Sea (capelin data from Gjøsæter et al., 2000, updated to 2003).


Figure 2.3.2.1 Annual variation in Partial Fullness Index (PFI) for age 0 cod from 1986-2002. Only years with more than 20 stomach samples are shown.

Age 1 cod


Figure 2.3.2.2. Annual variation in Partial Fullness Index (PFI) for age 1 cod from 1984-2002.

Age 2 cod


Figure 2.3.2.3. Annual variation in Partial Fullness Index (PFI) for age 2 cod from 1984-2002.

North east artic cod 0-group log index 1966-2005


Figure 2.3.4.1.1. The figure shows the 0 -group log index of North East Arctic cod (black) and the model fit (red), together with prognoses for 2004-2005 (green).

North East Artic cod, 3 year old,vpa2003,can,1983-2003


Figure 2.3.4.2.1. The figure shows the number of recruits (three year olds) of North East Arctic cod (black) and the model fit (red), together with prognoses for 2004-2005 (green).

Barents Sea Capelin
Recruits 1982-2003, with prognoses 2004


Figure 2.3.4.3.1. The figure shows the number of recruits (1 year olds) of Barents Sea Capelin (black) and the model fit (red), together with prognoses for 2004 (green).


Figure 2.3.4.4.1. The figure shows the number of recruits (3 year olds) of Norwegian spring spawning herring (black) and the model fit (red), together with prognoses for 2004-2006 (green).


Figure 2.4.1.1. Winter (December-March) North Atlantic Oscillation index (NAO). The 2004 value is calculated from December-February.


Figure 2.4.1.2 Temperature $\left({ }^{\circ} \mathrm{C}\right)$ and salinity observed during July/August, in the core of Atlantic Water beyond the shelf edge in the sections Svinøy - NW, Gimsøy - NW and Sørkapp - W, averaged between 50 and 200 m depth and horizontally over three stations across the core.



Figure 2.4.1.4 Time series of area (blue, in $\mathrm{km}^{2}$ ) and averaged temperature (red) of Atlantic water in the Svinøy section, observed in March/April (dotted line) and July/August (solid line) 1978-2004.

Ocean Weather Station Mike 2003


Figure 2.4.2.1 Distribution of chlorophyll $a$ at 10 m depth during the year at Weather Station Mike in 2003.


Figure 2.4.2.2 Year to year variations in the different phases of the development of phytoplankton at Weather Station Mike in the period 1991 to 2003. Diamonds: winter phase; squares: pre-bloom phase; triangles: spring bloom. Continuous lines represent the average for each phase. Broken lines represent one standard deviation for each phase.


Figure 2.4.3.1 Zooplankton biomass (dry weight) in the upper 200 m in May. A: Arctic influenced water (salinity $<35$, west of $1.4^{\circ} \mathrm{E}$ ). B: Atlantic water (salinity $>35$ ). B: Norwegian Coastal water (salinity $<35$, west of $1.4^{\circ} \mathrm{E}$ ). Error bars: $95 \%$ confidence limits.


Figure 2.4.3.2 Zooplankton biomass in July-August in the eastern Norwegian Sea ( $0-200 \mathrm{~m}$ ). Integrated biomass within a fixed geographical region divided by its area.

Herring condition index


Figure 2.4.4.1 Individual weight to length ratio (herring condition index) for Norwegian spring spawning herring. Data from November and December for herring $30-35 \mathrm{~cm}$ body length. Error bars: $95 \%$ confidence limits.


Figure 2.4.4.2 Zooplankton biomass (dry weight) in Atlantic water in the Norwegian Sea in May ( $0-200 \mathrm{~m}$ ) and herring condition index (individual weight to length ratio, November and December, 30-35 cm). Error bars: $95 \%$ confidence limits. Linear regression: Condition $=0.0045 *$ biomass $+0.7605 . \mathrm{R}^{2}=0.3434$.


Figure 2.4.5.1 Zooplankton biomass in July (year n) vs. zooplankton biomass in May (year n+1). Open triangles: 1996 and 2001.

March-April


Figure 2.4.5.2 Average North Atlantic oscillation index (NAO) during March and Aril (year n) vs. zooplankton biomass in May (year n+1). Circle: prediction of zooplankton biomass in May 2004 based on equation (1).

## March-April



Fig. 2.4.5.3 Herring condition index (year $n+1$ ) vs. average NAO during March and April (year n). Circle: prediction of herring condition in 2004 based on equation (2).


Figure 2.5.1.1 Temperature and Salinity deviations on the Siglunes section north of Icleand, mean for stations 1-5 and 0 - 200m, 1952-2003.


Figure 2.5.1.2 Variations in zooplankton biomass ( g dry weight $\mathrm{m}^{-2}, 0-50 \mathrm{~m}$ ) in spring at Siglunes section. The columns show means for 8 stations and the vertical bars denote standard error. The curved line shows 7 year running mean.


Figure 2.6 Yearly mean temperature and salinity from $50-600 \mathrm{~m}$ (crosses) of all stations in a box with bottom depth $>600 \mathrm{~m}$, west of the Porcupine bank bounded by $52^{\circ}$ to $54^{\circ} \mathrm{N}$ and 16 to $14^{\circ} \mathrm{W}$. Dotted lines are drawn at plus-minus one standard deviation of all observations in each box, each year.

### 3.1 General

### 3.1.1 Stock definition

The Norwegian spring spawning herring is a stock that is characterized by extended migrations, a high number of vertebrae, by large size at age, different scale characteristics from other herring stocks and large variation in year-class strength. The morphological characteristics of this stock resembled those of Icelandic spring spawning herring. However, the latter stock has disappeared.

ICES areas IIa, Ilb and I constitute the distribution area. The adult individuals of the Norwegian spring-spawning herring have a distinct annual migration pattern in the Norwegian Sea. This migration pattern changes over time, at present the herring winters in fjord areas in Northern Norway, spawn on the Norwegian coast (mainly between $62^{\circ}$ and $71^{\circ} \mathrm{N}$ ) and feed in the Norwegian Sea. The immature stock is distributed mainly in the Barents Sea, but some herring have their nursery area on the Norwegian coast. Historically, for instance in the period 1900-1950, significant spawning took place on the Norwegian coast south of $60^{\circ} \mathrm{N}$ (ICES area IV), but at present hardly any herring migrate to these spawning grounds. The adult component seldom mixes with herring of other stocks. However this can occur with immature herring (in certain fjord areas and in the eastern Barents Sea). In this case genetic characteristics are used as a supplement to the above mentioned separation criteria.

In the coming years special attention should be paid to the large 2002 year class, where a substantial part grew up in the Norwegian Sea instead of in the Barents Sea. It is expected that the growth of the Norwegian Sea part of this year class will be much higher than for the Barents Sea component, and also the natural mortality is expected to be lower in the Norwegian Sea component.

### 3.1.2 ACFM advice and management applicable to 2003 and 2004

In 2002 ACFM stated that "the stock is inside safe biological limits. The stock is harvested at or slightly below $\mathbf{F}_{\mathrm{pa}}=$ 0.15. The recruitment of the very strong 1992 year class led to an increase in SSB in 1997 to approximately 9 million t , but SSB has since declined to just over 5 million $t$ in 2001. The incoming year classes 1998 and 1999 are estimated to be strong." Further "ICES advises that this fishery should be managed according to the agreed management plan, corresponding to a catch of 710000 t in 2003".

In 2003 ACFM stated that "Based on the most recent estimate of SSB and fishing mortality ICES classifies the stock as being inside safe biological limits. The stock is harvested around $\mathbf{F}_{\mathrm{pa}}=0.15$. The recruitment of the very strong 1992 year class led to an increase in SSB in 1997 to approximately 8 million t , but SSB has since declined to just over 5 million t in 2002. The incoming year classes 1998 and 1999 are estimated to be relatively strong." Further "ICES advises that this fishery should be managed according to the agreed management plan with a fishing mortality of no more than $\mathrm{F}=0.125$, corresponding to landings in 2004 of less than 825000 t ."

There was no agreement on the TAC in 2003, but the sum of autonomous allocations from most of the individual Parties amount to 711500 tonnes.

At the meeting on Fisheries Consultation on the management of Norwegian spring-spawning herring (AtlantoScandian) herring stock in Reykjavik, Iceland in October 2003, the coastal states (European Union, Faroe Islands, Iceland, Norway, and Russia) did not reach any agreement regarding the allocation of the quota.

At a following meeting on Fisheries Consultation on the management of Norwegian spring-spawning herring stock in Copenhagen, Denmark, in mid February 2003, the parties were unable to reach any agreement on quota allocations. However, there seemed to be an unwritten understanding between the parties to accept the TAC proposed by ACFM to limit the total catches to 825000 t in 2004 .

### 3.1.3 Fishery

The total official catch in 2003 was 733.494 tonnes. The total catch used by the working group is 772.927 tonnes. The catches of Norwegian spring-spawning herring (NSSH) by all countries in 2003 by ICES rectangles are shown in Figure 3.1.3.1 (total whole year) and in Figure 3.1.3.2 (per quarter). In 2003 the catch provided as catch by rectangle represented $94.5 \%$ of the total catch. In general the development of the international fishery shown by these figures follows the known migration pattern for Norwegian spring-spawning herring. The migration pattern, together with
environmental factors, was mapped in 2003 during the ICES PGSPFN (Planning Group on Surveys on Pelagic Fish in the Norwegian Sea) investigations (ICES 2003/D:10 Ref ACFM, ACME).

Denmark: The Danish fishery of NSSH is carried out mostly by purse seiners and most of the catches were landed in Norway. In 2003 the first fishing period started in the southern part of Division IIa in February (catch 6,400 t) and continued into March when app. 6,100 t were caught. The second fishing period was in June in the Norwegian Sea and in the Jan Mayen area (app. 1,100 t). Landings from the latter half of the year were insignificant.

The Faroes: The Faroese fishery for NSSH (7 combined purse seiners/trawlers) started in late February 2003 in Norwegian EEZ (ICES Division IIa), relatively close to the coast off Sundmøre (Svinøy) and continued in that area and further north in early March. In mid May the fishery resumed in the international waters around $71-72^{\circ} \mathrm{N}$ (IIa), and simultaneously in the international area and Jan Mayen zone around $68-69^{\circ} \mathrm{N}$ and in the northern part of the Faroese area (IIa). The fishery continued in June and early July in the Jan Mayen area, gradually moving northeast into the international area and further into the Svalbard area (IIa and IIb) in late July and August. In late August the fishery had moved southeast into the Norwegian zone and continued in the Lofoten area and off Vesterålen ( $67-71^{\circ} \mathrm{N}$ ) in September and October. The fishery continued in the until late October. More than half of the catches were taken with pelagic trawls and the rest with purse-seines.

France: France reported no catches in 2003.
Germany: The main German fleet fishing for pelagic species is based in Bremerhaven and Rostock. It consists of 4 large pelagic freezer trawlers with overall lengths of $70 \mathrm{~m}, 95 \mathrm{~m}, 117 \mathrm{~m}$ and 124 m ; engine power $3700 \mathrm{hp}, 4400 \mathrm{hp}$, 8300 hp and 8600 hp and $25-40$ crew members. The vessels are owned and managed by Dutch companies but operating under German flag.

They are specially designed for pelagic fisheries: The catch is pumped into large refrigerated seawater tanks to keep the catch fresh until further processing, which is usually block freezing. The freezing capacity of the 4 vessels ranges between 60,000 and 230,000 cartons (up to $5,500 \mathrm{t}$ fish).

In 2003, the directed NSSH fishery started in early June with an exploratory fishery in rectangles 72F5 to 74F7. Only 179 t of herring were caught within 4 days. The next trip yielded 380 t in rect. 60 F 0 beginning of August. The bulk of German catch ( 2240 t ) was caught in the $2^{\text {nd }}$ and $3^{\text {rd }}$ week of August close to the border to division IIb (rect. 73 F 9 to 75 G 2 , and one haul in IIb). The last trip for herring in IIa was conducted in the $2^{\text {nd }}, 3^{\text {rd }}$ and $4^{\text {th }}$ week of September and yielded 277 t in rect. $73 \mathrm{~F} 9-\mathrm{G} 0,92 \mathrm{t}$ over the continental shelf break NW of the Lofoten and 164 t SW of the Lofoten.

Iceland: The Icelandic catch quota for NSSH was set at 103234 tonnes in 2003. The Icelandic fishery in 2003 began in the second week of May in international waters east of Jan Mayen and later in the month the fishery continued there as well as in the western part of the international waters in the Norwegian Sea between the Jan Mayen and Faroese zones (see the figure below). In June the fishery was mostly conducted in the international zone east-northeast of Jan Mayen but also at the borders of the Icelandic, Jan Mayen and international zones in the second week of June. In July the fishery moved still further north in the international zone and into the southern Spitzbergen zone and continued there untill the last week of August when it moved into the Norwegian EEZ. In September the fishery moved closer to the Norwegian coast and finished northwest of Lofoten in the last week of September. The bulk of the catch was caught in June ( 52 thous. tonnes) and May ( 26 thous. tonnes) and only about 1500 tonnes were caught in September. The total catch was 102737 tonnes of which 61882 tonnes were taken in midwater trawl and 40855 tonnes in purse-seine.

A total of 28 purse-seiners/trawlers participated in the herring fishery. The length range of the vessels was 47-79 meters with a mean length of 63 meters. The engine power range of the fleet was $441-5520 \mathrm{~kW}$ ( $599-7500 \mathrm{HP}$ ) with a mean of 2751 kW ( 3738 HP ).

Ireland: The information on the Ireland fishery was restricted to official catch.
Netherlands: The Dutch fleet fishing for pelagic species (herring, blue whiting, mackerel, horse mackerel, argentines) consists of 16 freezer trawlers and pair trawlers. They have a seasonal fishery in May-June directed towards AtlantoScandian herring in the Smutt-havet area (ICES Sub-area II) during which they also catch blue whiting.

Furthermore they have a seasonal fishery from February to April in ICES Divisions VIIb,c and VIa,b directed towards blue whiting.

Norway: The Norwegian fishery is carried out by many size categories of vessels. Of the total national quota of 484 $500 \mathrm{t}, 51 \%$ is allocated to purse seiners, $9 \%$ to trawlers and $39 \%$ to smaller coastal purse seiners. By far the larger part of the Norwegian fishery takes place in northern Norwegian coastal waters (Vestfjord area) where the herring winters from September until mid January. Here the herring occurs in concentrations that are easily available to the fishery. In 2003 only about 25000 tonnes were caught in the wintering areas in Northern Norway in January-February. 91000 t were taken in the spawning area on the Norwegian coast in February-March. Only negligible quantities ( 403 tonnes) were caught in the areas south of 62 N during the spawning period in 2003 . Only 160 t were caught in the spring/summer fishery in the Norwegian Sea. The remaining part of the Norwegian quota (approximately 297000 t ) was taken in the period September-December on the herring migrating to, and wintering in, the wintering areas in northern Norway. The total Norwegian catch in 2003 was 438140 tonnes

Poland: The Poland reported no catches in 2003.
Russia: In 2003 the Russian fishery started within the shelf region of the Norwegian EEZ, near Trena Bank (approximately $65^{\circ} \mathrm{N}-66^{\circ} \mathrm{N}$ ) in the beginning of February and Sklina and Buagrunnen Bank (approximately 63 ${ }^{\circ} \mathrm{N}-65^{\circ} \mathrm{N}$ ) in the end this month. In March the fishing was in progress in the same regions. In February and March the catch was 5054 t .

In May-June the commercial vessels conducted fishing in the northern part of the international area in the Norwegian Sea in region of the Polar Front and in the Jan Mayen area. In May-June the catch was 8262 t . In July-September vessels caught herring in the international area in the Norwegian Sea in the region of the Polar Front and the zone of Spitsbergen.

In September Russian vessels followed the southward migrating fish and continued their fishery in the Norwegian EEZ. In September the fishery of the herring was prolonged in the EEZ of Norway. The herring migrated southwestwards, along the depths of the continental slope. In July-September the catch was 68809 t . In the first decade of October the fishery of the herring were finished in the Norwegian EEZ. In October the catch was 18692 t . The total Russian catch of Norwegian spring spawning was 122846 tonnes of which about $5 \%$ were taken by purse seine and $95 \%$ with pelagic trawl. The entire Russian catch was utilized for human consumption.

Sweden: NSS herring is fished by seins and pelagic trawlers, in 2003 was around $60 \%$ landed by seins. The major part is landed in Norway. In the beginning of the year the catches stem from the southern parts of IIa and moves in the second half of the year further to the north.

UK (Scotland): The Scottish fishery for NSSH takes place at the end of the first quarter, once the mackerel season has finished and also at the end of the third quarter, once the North Sea herring fishery has been completed. Around one half of the pelagic fleet participates in this fishery using either single or pair trawl gear. Generally the catch is landed into factories in Norway and is worth around $£ 2.5$ million to the Scottish fleet.

### 3.2 Data

### 3.2.1 Commercial catch

The total annual catches of Norwegian spring-spawning herring for the period 1972-2003 (2003 preliminary) are presented in Tables 3.2.1.1 (by fishery) and 3.2.1.2 (by country).

The Working Group noted that in this fishery an unaccounted mortality caused by fishing operations and underreporting probably exists. In general, it was not possible to assess the magnitude of these extra removals from the stock, and taking into account the large catches taken in recent years, the relative importance of such additional mortality is probably low. Therefore, no extra amount to account for these factors has been added in 1994 and later years. In previous years, when the stock and the quotas were much smaller, an estimated amount of fish was added to the catches (Table 3.2.1.1).

In Røttingen et al (2002) the possibility of incorrect catch statistics because of the water content in herring deliveries and errors concerning conversion factors between fillets and live weight of herring were discussed. The paper gave an assessment of underreporting in 2001 of live weight of 18000 tonnes, which is $3 \%$ of the total catch of Norwegian spring-spawning herring landed in Denmark and Norway in 2001 (588 000 tonnes). This underreporting element is probably valid also for other pelagic stocks.

For 2003 Norway increased the allowed subtraction for water content from the weighted catches to $13 \%$. This value has been $4 \%$ in the years before 2003, thus an increase at $9 \%$. However, for 2004 EU and Norway have agreed to set the subtraction to $2 \%$. This percentage was set into practical effect from 1 February 2004.

As compared to earlier years the actual Norwegian catch thus rose by $9 \%$ in 2003 due to the application of a water content at $13 \%$. This accounts for a tonnage of 39433 which was added to the Norwegian catch for 2003 by the working group. This was done as the Norwegian catch is large and the additional tonnage is a significant contribution to the total catch and should be reflected in the catch numbers used by the working group. For 2004 (next report) the official catch will be used.

The combination of national catch-at-age and weight-at-age data for 2003 to obtain the total international catch-at-age and weight-at-age was done using the computer programme SALLOC, a standard ICES software. The official catch, sampled catch, and catch as used by the Working Group, together with number of samples, catch-at-age, and weight-atage for each fishery are given in Tables 3.2.1.3 and 3.2.1.4.

The Working Group noted that not all nations participating in the international fishery for Norwegian spring-spawning herring in 2003 had carried out an adequate sampling of their fishery. The allocation of catches for which no samples were taken and the final catch-at-age and weight-at-age by ICES area is given in Table 3.2.1.5. In general one used the Norwegian age distribution and weights for un-sampled fisheries in the Norwegian Sea in quarter 1,4, and the Russian age distributions and weight keys for quarter 3-4 for un-sampled fisheries in quarter 3-4. and the Iceland age distribution for quarter 2 for un-sampled fisheries in quarter 2. The Russian age distribution in quarter 3 was calculated using Russian length-age keys for quarter 3 for IIb ICES area.

### 3.2.2 Biological data

The natural mortalities and the weights in stock are in Table 3.2.2.1, weight in catches in Table 3.2.2.2, catches in Table 3.2.2.3, proportion maturity by age in Table 3.2.2.4 and natural mortality in Table 3.2.2.5. The weight in stock at 1 January 2004 was taken as the unweighted mean of Norwegian samples from December 2003for corresponding year classes, except for age 1 and 2 for which the mean weight of the corresponding year classes in the joint IMR-PINRO survey in the Barents Sea in September 2003 was used (Anon. 2003). The weight in catches in 2003 was taken from the total international weight-at-age (table 3.2.1.5) which were produced using the computer programme SALLOC, a standard ICES software.

### 3.2.3 Surveys

### 3.2.3.1 Spawning grounds

There was no acoustic survey to estimate the abundance of herring in the spawning areas in 2003. Earlier estimates are listed in Table 3.2.3.1.1.

### 3.2.3.2 Wintering areas

The wintering area of the herring is now split in two areas. This was first observed during the December survey in 2002 when concentrations of herring were observed by the RV Johan Hjort in oceanic waters off Vesterålen and Troms. In December 2003 a survey covering both the fjordic and oceanic wintering area was carried out by the RV G.O.Sars and R/V Johan Hjort. The combined result of the survey (Table 3.2.3.2.1) covers the entire known wintering area of the mature part of the stock. There was a very distinct difference in age structure between the two areas with most of the 1998 and 1999 yearclasses wintering in the oceanic areas and hardly any of the older yearclasses wintering in these areas.

An acoustic survey earlier took place on the wintering area in January. This series was ended in 1999 (Table 3.2.3.2.2).

### 3.2.3.3 Feeding areas

The feeding areas in the Norwegian Sea were surveyed acoustically by the ICES coordinated herring survey (PGNAPES, former PGPFN) during the period $25^{\text {th }}$ April to $10^{\text {th }}$ June 2003 (ICES 2003/D10). A complete set of the PGSPFN reports from the years 1995-2003 is found on www.imr.no/PGSPFN. The abundance estimate is given in Table 3.2.3.3.1.

### 3.2.3.4 Nursery area

The nursery areas of the Norwegian spring-spawning herring are Norwegian fjords and coastal areas, and in the Barents Sea. Since 1988, when the 1983-year class spawned for the first time, the latter area has increased in importance as a nursery area for the herring.

Results from the Norwegian and Russian acoustic survey in the Barents Sea in May-June 1991-2002 are given in Table 3.2.3.4.1. In 2001 the Working Group decided to include data on immature herring obtained during the RussianNorwegian survey in August-October in estimating the younger year classes in the Barents Sea. The results from these surveys are given in Table 3.2.3.4.2. The results from the 0 -group herring survey in Norwegian Fjords and Coastal areas are given in Table 3.2.3.4.3 and the results from the joint Norwegian-Russian 0-group survey in the Barents Sea are given in Table 3.2.3.4.4.

### 3.2.3.5 Herring larval survey

The larval survey in 2004 was carried out during the period 1-11 April. The survey started at Fugløya $\left(70^{\circ} \mathrm{N}\right)$ and continued along the Norwegian shelf south to Stad $\left(62^{\circ} \mathrm{N}\right)$. On a previous survey the area south of Stad had been covered, but no herring larvae were found in this area. High concentrations of herring larvae were found in the whole area north of Stad. Most of the larvae were in the first post yolksac stage and the mean length was 12.63 mm . This is more than 2 mm less than last year. The number of larvae was, however, much higher than last year (Table 3.2.3.5.1), and a higher number has only been recorded once, in 1997. Like the previous year a high proportion of the larvae were found in the northern part of the investigated area. The total number of larvae was estimated to be $56.4^{*} 10^{12}$.

### 3.2.4 Other relevant data

### 3.2.4.1 Tagging data

With the exception of 1999 and 2001, tagging has been carried out annually since 1975. The tagging experiments in 2003 were carried out in November in the wintering area (Tysfjorden, $68^{\circ} 05.5^{\prime} \mathrm{N}, 16^{\circ} 22.3^{\prime} \mathrm{E}$ ) where a total of 20798 herring were tagged. During the tagging process, the length of each tagged herring is measured. For each purse seine catch that is used for tagging, a sample of 100 fish is taken to determine the age distribution within each length group. The age composition of tagged herring in this batch is then estimated from the age distribution in the sample.

Recovery of tags from supervised detector plants has continued, as well as recovery from the standard magnets in the production line of fish processing plants and from individuals. For stock assessment purposes, tags are only used from supervised detector plants where detector efficiency has been tested, and where it is known that the detectors have been working as intended. Three factories filled these criteria in 2003, and a total of 88072 million herring were screened in these factories. Magnet efficiency was very close to $100 \%$ in 2003. All tagged herring recovered were sent to the Institute of Marine Research, Bergen, where they were measured, weighed and aged. In 2003130 tags from herring that were four years or more when tagged, were recovered from the factories (Table 3.2.4.1).

### 3.2.4.2 Prognosis of herring condition

A prognosis of the herring condition index for 2005 was used to estimate weight-at-age for the herring in 2005 in the short-term prognosis for NSSH. The prognosis of the herring condition was derived from a regression of herring condition during the time period 1990 to present on a two-month mean of the NAO index with a one-year time lag as described in Section 2.4.5. Figure 2.4 .3 shows the relationship between observed and modelled data of the herring condition.

### 3.3 Assessment models: Sea Star

In SeaStar (see 1.3.1) assessments he catchabilities of the acoustic surveys are assumed to have no age structure, i.e. no age or abundance dependence. The reason for this choice is that the mechanism for the recruiting of herring to the surveys is that of migration, not of growth, where the latter process could be assumed to be more regular and modellable. Instead, for each survey there is set a minimum age for inclusion of each year class. However, one needs to be cautious, since the different year classes may have different recruitment ages to the surveys. An important example is the failure of the 1998 year class to recruit fully to the December survey in 2002, for which reason the WG meeting 2003 considered this data point an outlier. In the coming years special attention should be paid also to the large 2002 year class, where a substantial part grew up in the Norwegian Sea instead of in the Barents Sea.

The input data used are:

Catch data
Acoustic surveys
Tag data
Larval data
0 -group data

Table 3.2.2.3
Table 3.2.3.1.1, 3.2.3.2.1, 3.2.3.3.1, 3.2.3.4.1, 3.2.3.4.2
Table 3.2.4.1
Table 3.2.3.5.1
Table 3.2.3.4.4

The acoustic surveys used are (the numbering is used elsewhere in the text of this section):

| 1 | Spawning grounds along the Norwegian coast | Minimum age: 5 |
| :--- | :--- | :--- |
| 2 | Wintering area in Vestfjorden in November-December | Minimum age: 4 |
| 3 | Wintering area in Vestfjorden in January | Minimum age: 5 |
| 4 | Young herring in the Barents Sea in May | Ages 1 and 2 |
| 5 | Feeding areas in the Norwegian Sea in May | Minimum age: 3 |
| 6 | As part of the joint IMR-PINRO capelin survey in <br> September | Ages 1 and 2 |

It is assumed that the distribution of the main tuning series of older fish follows a gamma distribution with a common CV, which is estimated, and that the distribution of the acoustic data in the Barents Sea follows a lognormal distribution. The tag return data are assumed to follow a Poisson distribution, which is commonly used for rare events, the larval data are assumed to follow a gamma distribution with an estimated CV and the zero group data are assumed to follow a normal distribution with an estimated standard deviation.

It has been experienced that when the fish get old the scales get difficult to read and more scales get discarded. This introduces a bias in the age distribution. An attempt has been made to corect for this age bias (WD by Schweder and Tjelmeland to the 2003 WG meeting).

Previously it has been observed in this WG that as the 1983 and 1985 year classes grew older than about 13 years the age readers tended to transfer fish from the 1983 to the 1985 year class. This problem is part of the general age reading problem for older fish, but is corrected for separately.

### 3.3.1 Comparison with the 2003 assessment

Before the exploratory analysis a comparison between the 2004 assessment and the 2003 assessment was made with the same settings as used in 2003, as recommended by ACFM. Figure 3.3.1.1 shows the perceived number by age for the two assessments together with the perceived number by age in 2003 in the assessment that later was chosen by the WG after an exploratory analysis.

With the same settings the new data do not lead to any appreciable change in the perceived stock in number by age in 2003. The 1998 year class has been slightly lifted. The run chosen by the WG as assessment this year leads to a lifting of the 1991 and 1998 year classes. The reason for this is that the data in the acoustic December and Norwegian Sea surveys, rather than information from the Barents Sea are now used to assess these year classes. For the same reason the 1997 year class is now perceived smaller. There is also a slight increase in the 1992 year class.

### 3.3.2 Data exploration with SeaStar

As has been the case during previous assessments of Norwegian spring spawning herring with SeaStar, a number of exploratory runs were initially performed to see the effect of various options and settings. These runs were:

| Label | Explanation |
| :--- | :--- |
| Default | With respect to the 2003 assessment the 1997-1999 year classes were <br> included in the tuning |
| The other runs are deviations <br> from Default: |  |
| NoTags | Tag data were left out |
| EstimateM | The natural mortality both for adult herring and young herring in the <br> Barents Sea were estimated |
| Oldest10 | Age 10 was set as the oldest true age group |
| CorrAgeBias | The bias in age reading was attempted corrected with the same method as |


|  | used in one exploratory run in 2002 |
| :--- | :--- |
| SurvSel | A survey selection curve was estimated along each tuned cohort |
| NoLarvae | The larval data were left out |

The estimated parameter catchabilities, spawning stock biomass in 2003 and 2002 and the contribution to the likelihood function from the various data sources are shown in Table 3.3.2.1 for the exploratory runs. Figure 3.3.2.1 shows the spawning stock timeseries for the runs. The catchabilities reflect how well each survey conforms to the modelled stock. The survey on the feeding areas in the Norwegian Sea seems to be the closest to the modelled stock, as has been the case also in previous assessments.

The perceived spawning stock is appreciably increased when the tagging data are removed and appreciably decreased when the larval data are removed. If $M$ is estimated there is a decrease in the spawning stock of about one million tonnes. The M -value for adult herring is estimated at 0.115 as opposed to 0.128 last year and the M -value for young herring in the Barents Sea is estimated at 0.188 , as opposed to 0.507 last year. The large reduction in perceived natural mortality for the young herring is probably connected to removing the 1997-1999 year classes from the Barents Sea estimation which resulted in an increase of year class strength for the 1998 and 1999 year class. If the oldest true age group is set to 10 years the spawning stock is drastically increased. However, this does also result in a drastic decrease of information. The trade-off between the benefit of including older fish in the analysis versus the problems connected to increasing bias as the fish grows old should be investigated further.

If the estimation allows for a survey selection to be estimated along each cohort the perceived spawning stock is increased. The change is, however, small, in accordance with the assumption made that once a year class is fully recruited to a survey the catchability remains constant.

### 3.4 Assessment models: ISVPA

### 3.4.1 Data exploration with ISVPA

For NSS herring exploratory runs by means of ISVPA the same version of the model and the same settings as in last year stock assessment were used (the catch-controlled version of the ISVPA with constraint of unbiased model approximation of logarithmic catch-at-age), except the following two aspects:

1. Starting year of analysis was 1950 instead of 1986 in previous years;
2. two selection patterns were fitted: first - for 1950-1985, and second - for 1986-2003.

Year of change in selection pattern was chosen not in correspondence with change in fishing regime, but later in order to retain similarity with previous assessments by means of ISVPA, when first year of analysis was 1986 and single selection pattern was estimated.

Profiles of components of ISVPA loss function for different sources of information are represented on figure 3.4.1.1. Survey in feeding area in May (N5 on figures) was treated the same way as in last year and was used for tuning in form of age proportions, weighted by stock abundance. This helps to exclude influence of possible occasional changes in this survey selection which may be the reason for deterioration of the signal from this survey when it is used for tuning in form of abundance-at-age. Weighting by abundance is used in order to make stress on abundant years as probably giving survey information of better quality. Treated as weighted age proportions, this survey corresponding to signals from most of other sources of information (compare curves N5 on figure 3.4.1.1 and figure 3.4.1.2, the last one representing signals for the case when survey N 5 is used in form of abundance-at-age). More considerations of survey data use in form of age proportions with examples of its implementation for real and simulated data are given in (Vasilyev, 2003). Overall loss function of the model (curve 7 on figure 3.4.1.1) has clear minimum.

If to restrict interval of years in the analysis to 1986-2003, the only change is slight change in the shape of the loss function component, corresponding to signal from catch-at-age data without significant change in its position (figure 3.4.1.3 to be compared to figure 3.4.1.1).

If to disregard all auxiliary information and find solution only on the basis of signal from catch-at-age alone, the result is very close to the result, when all information is used (including survey N5 in form of age proportions). This confirms that catch-at-age data, considered in frames of separability assumption, gives valuable signal about stock size and it seems to be not rational to disregard it. Results in terms of SSB for the above mentioned runs, as well as for runs, when age groups 3 and 4 were excluded from surveys N5, are presented on figure 3.4.1.4. For better resolution figure represents the estimates from 1990. As it can be seen, change in age diapason in survey 5 does not produce significant impact on results, while treatment of surveys 5 as abundance-at-age index shifts SSB estimates in final years down.

Bearing in mind that if survey N5 data are treated as age proportions (influence of possible changes in survey catchability is excluded), this data source gives the signal about the stock which is more in line with signals from most of other sources of information, including catch-at-age alone, in the final ISVPA run just this treatment of this data source was used. Other model settings in comparison with those used in previous assessment, are listed in the table 3.4.1.1.

Plots of residuals for all used sources of information are presented on figure 3.4.1.5. Looking at plot of residual in logarithmic catch-at-age it should be remind that for catch-controlled version, where abundance-at-age values are calculated via reported catch-at-age data and estimated selection pattern is used only for estimation of terminal abundance, residuals represent measure of separability, but not directly and not as measure of deviations of estimated abundances from "real" ones.

Figure 3.4.1.6 represents the estimated selection patterns (normalized tol by sum). The estimates show reasonably higher selection of younger ages in the first period.

Tables 3.4.1.2 and 3.4.1.3 represent estimates of abundance and fishing mortality coefficients. Summary results of final ISVPA run are presented in table 3.4.1.4 and on figure 3.4.1.7.

Results of retrospective runs are shown on figure 3.4.1.8. Estimates of fishing mortality are rather stable, while stability of estimates of recruitment is lower.

### 3.5 Exploratory assessments with ISVPA and SeaStar

The two methods (ISVPA, (WD Vasilyev 2004), and SeaStar Chapter 1.3) were tested, in an exploratory phase. The methods are different in the way that while the SeaStar method is tuned against abundance series only, and assuming that the catch data are without errors, the ISVPA is tuned against the abundance series and the catch data with the assumption of a constant selection pattern in the fisheries. In earlier years, the two methods have shown differences in results, both in terms of trends in the stock in recent years, but also in terms of stock levels. Last year, it was therefore not easy for the working group to agree on what method to use for the assessment.

### 3.5.1 Input data for comparative runs for testing of the methods

Exploratory assessments were made in order to compare the performance of the two methods. The tuning input data for these runs are in the text table below. The data sets were made as equal as possible, with almost the same abundance series. In SeaStar, it has been common to use tagging data. The tagging data has not been used in ISVPA, since these data has not been in the form of an abundance index series. The tagging data were therefore not applied in the runs for comparison of the methods. Both methods use, however, all the acoustic abundance estimates, with the same age ranges. In addition, ISVPA uses the catch data for tuning. SeaStar does not use the catch data for tuning, but in the back-calculation of the different year-classes (the VPA). The SeaStar method also uses the o-group series for the Barents Sea, a series which is not used in ISVPA. In addition, both methods use the standard input data for assessments such as stock weights, maturity proportion by year, natural mortalities etc.

Tuning data series in assessments
Catch at age data 1950-2003
Acoustic abundance estimates on the spawning grounds 1988-2000
Acoustic abundance series in the wintering areas in December 1992-2003
Acoustic abundance series in the wintering areas in January 1991-1999
Acoustic abundance series in the feeding area (Norwegian Sea) 1996-2002
Acoustic abundance series of young fish in the Barents Sea 1991-2003
Tagging experiments
O-group data from the Barents Sea 1965-2003
Larval surveys

## Series used for method test

X (ISVPA only)
X
X
X
X
X

X (SeaStar only)

### 3.5.2 Results

### 3.5.2.1 ISVPA

Figures 3.5.2.1 show the results of the minimisation of the object function of the different abundance indices and for the fit to the catch data for the ISVPA. The figures show well defined minima for all data series except for the young herring survey in the Barents Sea.

Figures 3.4.1.5 a-g show bubble plots of the residuals of the fit to the various data series.
The residuals (on log-scale) of the fit to the catch data, as shown in the bubble plots (Figure 3.4.1.5 a) have a tendency of high residuals for the young age groups ( 1 and 2 year olds). For the older age groups, the residuals are lower in value, but show a tendency of cohort effects.

The residual pattern of the fit to the survey indices (Figure 3.4.1.5 b-g) show somewhat higher residuals than those for the fit to the catch data, and there seem to be cohort effects. However, the higher values and the cohort effects seem to be associated with the small year-classes, especially the 1986 year-class.

### 3.5.2.2 SeaStar

Because the selection pattern in the fishery may vary according to the strength of the year- classes, the catch data are not used in the tuning procedures in SeaStar. In the recent 20 years, the 9 strongest cohorts are modelled and fitted to the survey data in a maximum likelihood procedure. The modelled year-class trajectories in relation to the observed once in acoustic estimates are shown in Figures 3.6.1-3.6.6. There is a variation in how well the model are fitted to the observed data. There seems to be both cohort and year effects for all series. The 1983 year-class seems to be overestimated in the wintering area surveys in January and on the spawning grounds, while the strong 1991 and 92 year-classes seem to be underestimated in the feeding area surveys.

### 3.5.3 The results of the two methods

The results in the form of spawning stock trajectories are shown in Figure 3.5.3. It shows that the methods performed similarly (showed the same overall trend and status of the stock) when the present set of data were used. However, it is at present impossible to compare the methods, in terms of statistical fit to the data because they do not give comparable statistical output.

The investigation of the two methods show that they give similar results with their present sets of data. The fulfilment of the assumption of the constant selection pattern is therefore shown not to be the major cause for the difference between the models. More important is the choice of data series and how these are handled.

### 3.5.4 Final runs

For more than ten years, tagging data has been used in the assessment of the Norwegian spring spawning herring. SeaStar can handle these data, and therefore when doing the assessment of the stock the tagging data were included as a tuning series. In this assessment, the larvae data are also included. The larvae data is an SSB index which has also been used for many years in the assessment. The result of this final SeaStar run is different from the one in the comparative runs and the results are given in Figure 3.5.3.

A final run with the ISVPA run was also made. For this method, additional data were not used, but the acoustic survey index for the Norwegian Sea was used for tuning of age proportions in the stock (detrended). This resulted in a difference in comparison with the comparative runs. The results are shown in Figure 3.5.3.

The working group chose the final run of the SeaStar method as the run to use as an assessment of the stock, and to use the results from this run as a basis for the predictions because this method incorporated all data series. However, SeaStar misses the information on stock abundance which may be found in the catch data and which ISVPA makes use of.

### 3.6 Final assessment including uncertainty and retrospective analysis

Figure 3.6 .1 shows the fit of the VPA scaled with the estimated catchability to the observations in the survey on the spawning grounds for the tuned year classes. Figure 3.6 .2 shows a similar fit for the December survey in Vestfjorden, Figure 3.6.3 for the January survey in Vestfjorden, Figure 3.6.4 for the May survey in the Barents Sea, Figure 3.6.5 for the feeding areas in the Norwegian Sea and Figure 3.6.6 for the Barents Sea September survey. With exception of the 1991 year class in the December survey, possibly the 1983 year class in the survey on the spawning grounds and the 1999 and 1998 year class in the survey on the feeding grounds all year classes seem to recruit well to all surveys and have a reasonably good fit. However, the 1996 year class seems to be somewhat systematically underrepresented in the survey on the feeding areas.

Table 3.6.1 shows the modelled historic abundance and Table 3.6.2 the F-values for the period 1950-2003. Table 3.6.3 shows the historic development of spawning stock, recruitment and abundance-averaged F -values over ages 5-14. It should be noted that SeaStar calculates the number of fish in the plus-group by summing the number of fish in the plusgroup the year before and the number of fish in the oldest true age group the year before, adjusted for natural mortality and catch. Therefore, the spawning stock time series might not be comparable with data calculated earlier with a different method.

Figure 3.6 .7 shows the retrospective plot. The assessment the present year is well in accordance with the assessment made in 2003. In 1999 there was a substantial increase in the assessment from the preceding years. It is interesting to note that in 1999 all year classes increased with respect to the previous year in the January survey and the survey on the spawning grounds, as all year classes also did in the September survey in 1998. These increases had a substantial effect on the assessment, but this effect diminished in the following years.

Figure 3.6 .8 shows a histogram of the spawning stock in 2004 for the bootstrap replicates. The standard deviation is about 4 million tonnes - a drastic increase from last year. The distribution is also more skewed than last year. This reason for the changes of the distributional form should be investigated further. However, the median spawning stock is close to the maximum likelihood value.

Figure 3.6.9 shows a scatter plot of the co-variation between terminal F-values and the spawning stock. While the covariation between the terminal F-values for the adult year classes and the spawning stock show a smooth distribution, the scatter plot involving the terminal F-values for the 2002 and 2003 year classes seems to consist of disjunctive clusters. This might be an indication of a problem in using the Barents Sea data which should be looked upon more closely.

Figure 3.6.10 shows the spawning stock time series for the present assessment together with the spawning stock time series for the 2003 assessment, Figure 3.6.11 a similar plot for the abundance averaged F for ages 5-14 and Figure 3.6.12 a similar plot for recruitment at age 0 .

The main settings for the final SeaStar runs in 2003 and 2004 are found in table 3.6.4.

### 3.6.1 Short term projection

Table 3.6.1.1 shows the input data to the short term prediction. For the weight at age in 2005 the weight at age in 2004 which is taken from December 2003 data was multiplied with the ratio of the projected condition factor in December 2004 to the measured condition factor in December 2003. The condition factors are described in section 3.2.4.2. The exploitation pattern was set to the F -values estimated for 2003. It is assumed that a catch of 0.825 million tonnes is taken in 2004. The weight at age in the catch is taken as the weight at age in the catch in 2003 and adjusted with the projected condition factor as is the case for the weight at age in the stock. The proportion mature at age was set equal to the proportion mature at age in 2003 with the exception for an adjustment for the 2002 year class as 3 year old which as set at 0.15 instead of 0.0 .

Table 3.6.1.2 shows the results of the short term prediction. For a population weighted fishing mortality over the ages 514 of about the agreed value of 0.125 the expected catch in 2004 is about 0.9 million tonnes. The spawning stock will remain virtually unchanged from 2004 to 2006.

### 3.7 Biological reference points

The process of establishing biological reference points (including a $\mathbf{B}_{\text {lim }}$ a value of 2.5 million tonnes) for this stock is given in Røttingen (2000). The SGPRP03 stated that the segmented regression was significant on a $5 \%$ level and gave a change point of 2.3 million tonnes. Further, the SGPRP03 has the opinion that these numbers are close to each other and as this stock is managed to an agreed harvest control rule, there is no need to change the $\mathbf{B}_{\mathrm{lim}}$ value.

If the spawning stock falls below 5.0 million tonnes $\left(\mathbf{B}_{\mathrm{pa}}\right)$ the managers have, according to the harvest control rule, decided on a linear reduction in fishing mortality from the agreed maximum fishing mortality 0.125 ( $\mathbf{F}_{\mathrm{pa}}$ is set to 0.15 ) at $\mathbf{B}_{\mathrm{pa}}$ to 0.05 at $\mathbf{B}_{\text {lim }}$. The WGNPBW agrees on the conclusion of the SGPRP03 on the $\mathbf{B}_{\mathrm{lim}}$, and is, due to the elements in the agreed harvest control rule, also of the opinion that there is no reason at present to change the values for the biological reference points $\mathbf{F}_{\mathrm{pa}}$ and $\mathbf{B}_{\mathrm{pa}}$.

EU, Faroe Islands, Iceland, Norway, and Russia agreed to implement a long-term management plan. This plan consists of the following elements (ICES 2002/CRR:255):

1. Every effort shall be made to maintain a level of Spawning Stock Biomass (SSB) greater than the critical level ( $\boldsymbol{B}_{\text {lim }}$ ) of $2500000 t$.
2. For the year 2001 and subsequent years, the Parties agreed to restrict their fishing on the basis of a TAC consistent with a fishing mortality rate of less than 0.125 for appropriate age groups as defined by ICES, unless future scientific advice requires modification of this fishing mortality rate.
3. Should the SSB fall below a reference point of $5000000 t\left(\boldsymbol{B}_{p a}\right)$, the fishing mortality rate, referred under paragraph 2, shall be adapted in the light of scientific estimates of the conditions to ensure a safe and rapid recovery of the SSB to a level in excess of 5000000 t. The basis for such an adaptation should be at least a linear reduction in the fishing mortality rate from 0.125 at $\boldsymbol{B}_{p a}(5000000 t)$ to $0.05 \boldsymbol{B}_{\text {lim }}(2500000 t$ ).
4. The Parties shall, as appropriate, review and revise these management measures and strategies on the basis of any new advice provided by ICES.

ICES considers that the objectives of this agreement are consistent with the precautionary approach.

Table 3.2.1.1 Catches of Norwegian spring-spawning herring (tonnes) since 1972.

| Year | A | $\mathrm{B}^{1}$ | C | D | Total | Total catch <br> used in WG |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 1972 | - | 9895 | $3,266^{2}$ | - | 13,161 | 13,161 |
| 1973 | 139 | 6,602 | 276 | - | 7,017 | 7,017 |
| 1974 | 906 | 6,093 | 620 | - | 7,619 | 7,619 |
| 1975 | 53 | 3,372 | 288 | - | 3,713 | 13,713 |
| 1976 | - | 247 | 189 | - | 436 | 10,436 |
| 1977 | 374 | 11,834 | 498 | - | 12,706 | 22,706 |
| 1978 | 484 | 9,151 | 189 | - | 9,824 | 19,824 |
| 1979 | 691 | 1,866 | 307 | - | 2,864 | 12,864 |
| 1980 | 878 | 7,634 | 65 | - | 8,577 | 18,577 |
| 1981 | 844 | 7,814 | 78 | - | 8,736 | 13,736 |
| 1982 | 983 | 10,447 | 225 | - | 11,655 | 16,655 |
| 1983 | 3,857 | 13,290 | 907 | - | 18,054 | 23,054 |
| 1984 | 18,730 | 29,463 | 339 | - | 48,532 | 53,532 |
| 1985 | 29,363 | 37,187 | 197 | 4,300 | 71,047 | 169,872 |
| 1986 | $71,122^{3}$ | 55,507 | 156 | - | 126,785 | 225,256 |
| 1987 | 62,910 | 49,798 | 181 | - | 112,899 | 127,306 |
| 1988 | 78,592 | 46,582 | 127 | - | 125,301 | 135,301 |
| 1989 | 52,003 | 41,770 | 57 | - | 93,830 | 103,830 |
| 1990 | 48,633 | 29,770 | 8 | - | 78,411 | 86,411 |
| 1991 | 48,353 | 31,280 | 50 | - | 79,683 | 84,683 |
| 1992 | 43,688 | 55,737 | 23 | - | 99,448 | 104,448 |
| 1993 | 11,195 | 110,212 | 50 | - | 227,457 | 232,457 |
| 1994 | 288,581 | 190,643 | 4 | - | 479,228 | 479,228 |
| 1995 | 320,731 | 581,495 | 0 | - | 902,226 | 902,226 |
| 1996 | 462,248 | 758,035 |  | 0 | - | $1,220,283$ |
| $1997^{5}$ |  |  | 0 | - | $1,426,507$ | $1,220,283$ |
| $1998^{5}$ |  |  | 0 | - | $1,223,131$ | $1,223,507$ |
| $1999^{6}$ |  |  |  | 0 | - | $1,235,433$ |
| $2000^{7}$ |  |  |  | 0 | - | $1,207,201$ |

$\mathrm{A}=$ catches of adult herring in winter
$\mathrm{B}=$ mixed herring fishery in remaining part of the year
$\mathrm{C}=$ by-catches of 0 - and 1-group herring in the sprat fishery
$\mathrm{D}=$ USSR-Norway by-catch in the capelin fishery (2-group)
Includes also by-catches of adult herring in other fisheries In 1972, there was also a directed herring 0 -group fishery
Includes $26,000 t$ of immature herring (1983 year class) fished by USSR in the Barents Sea Preliminary, as provided by Working Group members
5 Details of catches by fishery and ICES area given in ICES 1999
${ }_{7}$ Details of catches by fishery and ICES area given in ICES 2000
${ }^{7}$ Details of catches by fishery and ICES area given in ICES 2001
${ }^{8}$ Details of catches by fishery and ICES area given in ICES 2002
${ }^{9}$ Details of catches by fishery and ICES area given in ICES 2003
${ }^{10}$ Details of catches by fishery and ICES area given in Tables 3.2.3-3.2.5

Table 3.2.1.2 Total catch of Norwegian spring-spawning herring (tonnes) since 1972. Data provided by Working Group members.

| Year | Norway | USSR/ | Denmar | Faroes | Iceland | Ireland | Nether- | Greenl |  | German | Franc | Poland | Swede | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Russia | k |  |  |  | lands | and |  | y | e |  | n |  |
| 1972 | 13,161 | - | - | - | - | - | - | - | - |  | - |  | - | 13,161 |
| 1973 | 7,017 | - | - | - | - | - | - | - | - |  | - |  | - | 7,017 |
| 1974 | 7,619 | - | - | - | - | - | - | - | - |  | - |  | - | 7,619 |
| 1975 | 13,713 | - | - | - | - | - |  | - | - |  | - |  | - | 13,713 |
| 1976 | 10,436 | - | - | - | - | - | - | - | - |  | - |  | - | 10,436 |
| 1977 | 22,706 | - | - | - | - | - | - | - | - |  | - |  | - | 22,706 |
| 1978 | 19,824 | - | - | - | - | - | - | - | - |  | - |  | - | 19,824 |
| 1979 | 12,864 | - | - | - | - | - | - | - | - |  | - |  | - | 12,864 |
| 1980 | 18,577 | - | - | - | - | - | - | - | - |  | - |  | - | 18,577 |
| 1981 | 13,736 | - | - | - | - | - | - | - | - |  | - |  | - | 13,736 |
| 1982 | 16,655 | - | - | - | - | - | - | - | - |  | - |  | - | 16,655 |
| 1983 | 23,054 | - | - | - | - | - | - | - | - |  | - |  | - | 23,054 |
| 1984 | 53,532 | - | - | - | - | - | - | - | - |  | - |  | - | 53,532 |
| 1985 | 167,272 | 2,600 | - | - | - | - | - | - | - |  | - |  | - | 169,872 |
| 1986 | 199,256 | 26,000 | - | - | - | - | - | - | - | - | - |  | - | 225,256 |
| 1987 | 108,417 | 18,889 | - | - | - | - | - | - | - |  | - |  | - | 127,306 |
| 1988 | 115,076 | 20,225 | - | - | - | - | - | - | - |  | - |  | - | 135,301 |
| 1989 | 88,707 | 15,123 | - | - | - | - | - | - | - |  | - |  | - | 103,830 |
| 1990 | 74,604 | 11,807 | - | - | - | - | - | - | - |  | - |  | - | 86,411 |
| 1991 | 73,683 | 11,000 | - | - | - | - | - | - | - |  | - |  | - | 84,683 |
| 1992 | 91,111 | 13,337 | - | - | - | - | - | - | - |  | - |  | - | 104,448 |
| 1993 | 199,771 | 32,645 | - | - | - | - | - | - | - | - | - |  | - | 232,457 |
| 1994 | 380,771 | 74,400 | - | 2,911 | 21,146 | - | - | - | - | - | - |  | - | 479,228 |
| 1995 | 529,838 | 101,987 | 30,577 | 57,084 | 174,109 | - | 7,969 | 2,500 | 881 | 556 | - |  | - | 905,501 |
| 1996 | 699,161 | 119,290 | 60,681 | 52,788 | 164,957 | 19,541 | 19,664 | - | 46,131 | 11,978 | - |  | 22,424 | 1,220,283 |
| 1997 | 860,963 | 168,900 | 44,292 | 59,987 | 220,154 | 11,179 | 8,694 |  | 25,149 | 6,190 | 1,500 |  | -19,499 | 1,426,507 |
| 1998 | 743,925 | 124,049 | 35,519 | 68,136 | 197,789 | 2,437 | 12,827 |  | 15,9711 | 7,003 | 605 |  | -14,863 | 1,223,131 |
| 1999 | 740,640 | 157,328 | 37,010 | 55,527 | 203,381 | 2,412 | 5,871 | - | 9,207 | - | - |  | 14,057 | 1,235,433 |
| 2000 | 713,500 | 163,261 | 34,968 | 68,625 | 186,035 | 8,939 | - | - | 14,096 | 3,298 | - |  | 14,749 | 1,207,201 |
| 2001 | 495,036 | 109,054 | 24,038 | 34,170 | 77,693 | + | 6,439 | - | 12,230 | 1,588 | - |  | 9,818 | 770,066 |
| 2002 | 487,233 | 113,763 | 18,998 | 32,302 | 127,197 | + | 9,392 | - | 3,482 | 3,017 | - | 1,226 | 9,486 | 806,086 |
| $2003{ }^{1}$ | 438,140 | 122,846 | 14,144 | 27,943 | 102,727 | + | 8,678 | - | 9,214 | 3,371 | - | - | 6,431 | 733,494 |

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Table 3.2.1.5
Summary of Sampling by Country - Norwegian spring spawning herring.
$\qquad$
AREA : IVa


AREA : IIb


AREA : IIa

| Country | Sampled Catch | Official <br> Catch | No. of samples | No. <br> measured | No. aged | $\begin{gathered} \text { SOP } \\ \% \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK (Scot) | 0.00 | 9214.00 | 0 | 0 | 0 | 0.00 |
| Sweden | 0.00 | 6431.00 | 0 | 0 | 0 | 0.00 |
| Russia | 82135.00 | 100827.00 | 91 | 11828 | 2629 | 98.72 |
| Norway | 374353.00 | 477125.00 | 266 | 21451 | 7521 | 99.79 |
| Netherlands | 0.00 | 8678.00 | 0 | 0 | 0 | 0.00 |
| Iceland | 78051.00 | 93875.00 | 16 | 700 | 654 | 144.99 |
| Germany | 0.00 | 3307.00 | 0 | 0 | 0 | 0.00 |
| Faroes | 0.00 | 26349.00 | 0 | 0 | 0 | 0.00 |
| Denmark | 13237.00 | 14144.00 | 7 | 732 | 728 | 100.01 |
| Total IIa | 547776.00 | 739950.00 | 380 | 34711 | 11532 | 106.08 |
| Sum of Of | ches : | 739950.00 |  |  |  |  |
| Unallocat |  | 0.00 |  |  |  |  |
| Working G | h : | 739950.00 |  |  |  |  |

Table 3.2.1.5 continued

PERIOD : 1

| Country | Sampled Catch | Official Catch |
| :---: | :---: | :---: |
| UK (Scot) | 0.00 | 5157.00 |
| Sweden | 0.00 | 1695.00 |
| Russia | 5064.00 | 5064.00 |
| Norway | 128918.00 | 140586.00 |
| Faroes | 0.00 | 1416.00 |
| Denmark | 12127.00 | 12127.00 |
| Period Total | 146109.00 | 166045.00 |
| Sum of Offical Catches : Unallocated Catch : |  | 166045.00 |
|  |  | 0.00 |


| No. of | No. |
| :---: | :---: |
| samples | measured |
| 0 | 0 |
| 0 | 0 |
| 57 | 6876 |
| 111 | 11387 |
| 0 | 0 |
| 4 | 487 |
| 172 | 18750 |


| No. | SOP |
| :---: | ---: |
| aged | $\%$ |
| 0 | 0.00 |
| 0 | 0.00 |
| 2019 | 99.86 |
| 1812 | 99.80 |
| 0 | 0.00 |
| 483 | 100.02 |
| 4314 | 99.82 |

PERIOD : 2

| Country | Sampled Catch | Official Catch |
| :---: | :---: | :---: |
| Russia | 8262.00 | 8262.00 |
| Norway | 0.00 | 6671.00 |
| Netherlands | 0.00 | 6809.00 |
| Iceland | 78051.00 | 78712.00 |
| Germany | 0.00 | 180.00 |
| Faroes | 0.00 | 7608.00 |
| Denmark | 1110.00 | 1110.00 |
| Period Total | 87423.00 | 109352.00 |
| Sum of Offical Catches :Unallocated Catch : |  | 109352.00 |
|  |  | 0.00 |
| Unallocated Catch : |  | 109352.00 |


| No. of | No. |
| :---: | :---: |
| samples | measured |
| 11 | 1402 |
| 0 | 0 |
| 0 | 0 |
| 16 | 700 |
| 0 | 0 |
| 0 | 0 |
| 3 | 245 |
| 30 | 2347 |


| No. | SOP |
| ---: | ---: |
| aged | $\%$ |
| 250 | 99.89 |
| 0 | 0.00 |
| 0 | 0.00 |
| 654 | 144.99 |
| 0 | 0.00 |
| 0 | 0.00 |
| 245 | 99.89 |
| 1149 | 140.16 |

PERIOD : 3

| Country | Sampled <br> Catch | Official <br> Catch |
| :--- | ---: | ---: |
| UK(Scot) | 0.00 | 4057.00 |
| Sweden | 0.00 | 1950.00 |
| Russia | 68809.00 | 90828.00 |
| Norway | 0.00 | 62783.00 |
| Netherlands | 0.00 | 1869.00 |
| Iceland | 0.00 | 24015.00 |
| Germany | 0.00 | 3191.00 |
| Faroes | 0.00 | 13407.00 |
| Denmark | 0.00 | 268.00 |
| $\quad$ Period Total | 68809.00 | 202368.00 |
| $\quad$ Sum of Offical Catches | $:$ | 202368.00 |
| Unallocated Catch : |  | 0.00 |
| Working Group Catch |  |  |

PERIOD : 4

| Country | Sampled Catch | Official Catch | No. of samples | No. measured | No. aged | $\begin{gathered} \text { SOP } \\ \% \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sweden | 0.00 | 2786.00 | 0 | 0 | 0 | 0.00 |
| Russia | 0.00 | 18692.00 | 0 | 0 | 0 | 0.00 |
| Norway | 245435.00 | 267533.00 | 155 | 10064 | 5709 | 99.78 |
| Faroes | 0.00 | 5512.00 | 0 | 0 | 0 | 0.00 |
| Denmark | 0.00 | 639.00 | 0 | 0 | 0 | 0.00 |
| Period Total | 245435.00 | 295162.00 | 155 | 10064 | 5709 | 99.78 |
| Sum of Offical | ches : | 295162.00 |  |  |  |  |
| Unallocated Cat |  | 0.00 |  |  |  |  |
| Working Group C | h : | 295162.00 |  |  |  |  |

Table 3.2.1.5 continued
Total over all Areas and Periods

| Country | Sampled Catch | $\begin{gathered} \text { Official } \\ \text { Catch } \end{gathered}$ | No. of samples | $\begin{aligned} & \text { No. } \\ & \text { measured } \end{aligned}$ | No. aged | $\begin{gathered} \text { SOP } \\ \% \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK (Scot) | 0.00 | 9214.00 | 0 | 0 | 0 | 0.00 |
| Sweden | 0.00 | 6431.00 | 0 | 0 | 0 | 0.00 |
| Russia | 82135.00 | 122846.00 | 91 | 11828 | 2629 | 98.72 |
| Norway | 374353.00 | 477573.00 | 266 | 21451 | 7521 | 99.79 |
| Netherlands | 0.00 | 8678.00 | 0 | 0 | 0 | 0.00 |
| Iceland | 78051.00 | 102727.00 | 16 | 700 | 654 | 144.99 |
| Germany | 0.00 | 3371.00 | 0 | 0 | 0 | 0.00 |
| Faroes | 0.00 | 27943.00 | 0 | 0 | 0 | 0.00 |
| Denmark | 13237.00 | 14144.00 | 7 | 732 | 728 | 100.01 |
| Total for Stock | 547776.00 | 772927.00 | 380 | 34711 | 11532 | 106.08 |
| Sum of Offical c | ches : | 772927.00 |  |  |  |  |
| Unallocated Catch |  | 0.00 |  |  |  |  |
| Working Group Ca | h : | 772927.00 |  |  |  |  |

## DETAILS OF DATA FILLING-IN



## Table 3.2.1.5 continued


>> (10) Russia 3 IIa

## Table 3.2.1.5 continued

Catch Numbers at Age by Area

| Ages | IVa | IIb | IIa | Total |
| :---: | :---: | :---: | :--- | :--- |
| 0 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1 | 0.00 | 273.25 | 3163.60 | 3436.84 |
| 2 | 0.00 | 244.88 | 4257.12 | 4502.00 |
| 3 | 12.69 | 11109.47 | 63403.38 | 74525.54 |
| 4 | 70.81 | 27019.18 | 299133.03 | 326223.00 |
| 5 | 260.45 | 36649.90 | 713443.94 | 750354.25 |
| 6 | 82.23 | 5503.21 | 176088.75 | 181674.19 |
| 7 | 122.20 | 4105.78 | 168598.55 | 172826.53 |
| 8 | 6.19 | 1092.98 | 22873.35 | 23972.52 |
| 9 | 32.19 | 3727.11 | 73680.24 | 77439.54 |
| 10 | 100.54 | 10314.32 | 211805.73 | 222220.59 |
| 11 | 664.08 | 9864.26 | 567839.50 | 578367.88 |
| 12 | 151.22 | 2888.46 | 220107.83 | 223147.52 |
| 13 | 21.84 | 779.55 | 43258.66 | 44060.05 |
| 14 | 8.86 | 0.00 | 9109.77 | 9118.64 |
| 15 | 5.80 | 0.00 | 4760.62 | 4766.42 |

Mean Weight at Age by Area (Kg)

| Ages |  |  |  | IVa |
| :---: | :---: | :---: | :---: | :---: |
| IIb | IIa |  | Total |  |
| 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1 | 0.0000 | 0.0850 | 0.0598 | 0.0618 |
| 2 | 0.0000 | 0.1270 | 0.0646 | 0.0680 |
| 3 | 0.1516 | 0.1690 | 0.1696 | 0.1695 |
| 4 | 0.1833 | 0.2170 | 0.2157 | 0.2158 |
| 5 | 0.2256 | 0.2660 | 0.2543 | 0.2549 |
| 6 | 0.2552 | 0.3130 | 0.2846 | 0.2855 |
| 7 | 0.2800 | 0.3480 | 0.3160 | 0.3167 |
| 8 | 0.3161 | 0.3270 | 0.3251 | 0.3252 |
| 9 | 0.3167 | 0.3840 | 0.3489 | 0.3506 |
| 10 | 0.3102 | 0.3940 | 0.3585 | 0.3602 |
| 11 | 0.3176 | 0.4020 | 0.3573 | 0.3581 |
| 12 | 0.3369 | 0.4050 | 0.3781 | 0.3784 |
| 13 | 0.3495 | 0.4590 | 0.3931 | 0.3942 |
| 14 | 0.3329 | 0.0000 | 0.3994 | 0.3994 |
| 15 | 0.4313 | 0.0000 | 0.4476 | 0.4476 |

Table 3.2.2.1 Weight at age in the stock for Norwegian spring spawning herring, gram

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 1 | 8 | 47 | 100 | 204 | 230 | 255 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1951 | 1 | 8 | 47 | 100 | 204 | 230 | 255 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1952 | 1 | 8 | 47 | 100 | 204 | 230 | 255 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1953 | 1 | 8 | 47 | 100 | 204 | 230 | 255 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1954 | 1 | 8 | 47 | 100 | 204 | 230 | 255 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1955 | 1 | 8 | 47 | 100 | 195 | 213 | 260 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1956 | 1 | 8 | 47 | 100 | 205 | 230 | 249 | 275 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1957 | 1 | 8 | 47 | 100 | 136 | 228 | 255 | 262 | 290 | 305 | 315 | 325 | 330 | 340 | 345 | 362 | 365 |
| 1958 | 1 | 8 | 47 | 100 | 204 | 242 | 292 | 295 | 293 | 305 | 315 | 330 | 340 | 345 | 352 | 360 | 365 |
| 1959 | 1 | 8 | 47 | 100 | 204 | 252 | 260 | 290 | 300 | 305 | 315 | 325 | 330 | 340 | 345 | 355 | 360 |
| 1960 | 1 | 8 | 47 | 100 | 204 | 270 | 291 | 293 | 321 | 318 | 320 | 344 | 349 | 370 | 379 | 375 | 380 |
| 1961 | 1 | 8 | 47 | 100 | 232 | 250 | 292 | 302 | 304 | 323 | 322 | 321 | 344 | 357 | 363 | 365 | 370 |
| 1962 | 1 | 8 | 47 | 100 | 219 | 291 | 300 | 316 | 324 | 326 | 335 | 338 | 334 | 347 | 354 | 358 | 358 |
| 1963 | 1 | 8 | 47 | 100 | 185 | 253 | 294 | 312 | 329 | 327 | 334 | 341 | 349 | 341 | 358 | 375 | 375 |
| 1964 | 1 | 8 | 47 | 100 | 194 | 213 | 264 | 317 | 363 | 353 | 349 | 354 | 357 | 359 | 365 | 402 | 402 |
| 1965 | 1 | 8 | 47 | 100 | 186 | 199 | 236 | 260 | 363 | 350 | 370 | 360 | 378 | 387 | 390 | 394 | 394 |
| 1966 | 1 | 8 | 47 | 100 | 185 | 219 | 222 | 249 | 306 | 354 | 377 | 391 | 379 | 378 | 361 | 383 | 383 |
| 1967 | 1 | 8 | 47 | 100 | 180 | 228 | 269 | 270 | 294 | 324 | 420 | 430 | 366 | 368 | 433 | 414 | 414 |
| 1968 | 1 | 8 | 47 | 100 | 115 | 206 | 266 | 275 | 274 | 285 | 350 | 325 | 363 | 408 | 388 | 378 | 378 |
| 1969 | 1 | 8 | 47 | 100 | 115 | 145 | 270 | 300 | 306 | 308 | 318 | 340 | 368 | 360 | 393 | 397 | 397 |
| 1970 | 1 | 8 | 47 | 100 | 209 | 272 | 230 | 295 | 317 | 323 | 325 | 329 | 380 | 370 | 380 | 391 | 391 |
| 1971 | 1 | 15 | 80 | 100 | 190 | 225 | 250 | 275 | 290 | 310 | 325 | 335 | 345 | 355 | 365 | 390 | 390 |
| 1972 | 1 | 10 | 70 | 150 | 150 | 140 | 210 | 240 | 270 | 300 | 325 | 335 | 345 | 355 | 365 | 390 | 390 |
| 1973 | 1 | 10 | 85 | 170 | 259 | 342 | 384 | 409 | 404 | 461 | 520 | 534 | 500 | 500 | 500 | 500 | 500 |
| 1974 | 1 | 10 | 85 | 170 | 259 | 342 | 384 | 409 | 444 | 461 | 520 | 543 | 482 | 482 | 482 | 482 | 482 |
| 1975 | 1 | 10 | 85 | 181 | 259 | 342 | 384 | 409 | 444 | 461 | 520 | 543 | 482 | 482 | 482 | 482 | 482 |
| 1976 | 1 | 10 | 85 | 181 | 259 | 342 | 384 | 409 | 444 | 461 | 520 | 543 | 482 | 482 | 482 | 482 | 482 |
| 1977 | 1 | 10 | 85 | 181 | 259 | 343 | 384 | 409 | 444 | 461 | 520 | 543 | 482 | 482 | 482 | 482 | 482 |
| 1978 | 1 | 10 | 85 | 180 | 294 | 326 | 371 | 409 | 461 | 476 | 520 | 543 | 500 | 500 | 500 | 500 | 500 |
| 1979 | 1 | 10 | 85 | 178 | 232 | 359 | 385 | 420 | 444 | 505 | 520 | 551 | 500 | 500 | 500 | 500 | 500 |
| 1980 | 1 | 10 | 85 | 175 | 283 | 347 | 402 | 421 | 465 | 465 | 520 | 534 | 500 | 500 | 500 | 500 | 500 |
| 1981 | 1 | 10 | 85 | 170 | 224 | 336 | 378 | 387 | 408 | 397 | 520 | 543 | 512 | 512 | 512 | 512 | 512 |
| 1982 | 1 | 10 | 85 | 170 | 204 | 303 | 355 | 383 | 395 | 413 | 453 | 468 | 506 | 506 | 506 | 506 | 506 |
| 1983 | 1 | 10 | 85 | 155 | 249 | 304 | 368 | 404 | 424 | 437 | 436 | 493 | 495 | 495 | 495 | 495 | 495 |
| 1984 | 1 | 10 | 85 | 140 | 204 | 295 | 338 | 376 | 395 | 407 | 413 | 422 | 437 | 437 | 437 | 437 | 437 |
| 1985 | 1 | 10 | 85 | 148 | 234 | 265 | 312 | 346 | 370 | 395 | 397 | 428 | 428 | 428 | 428 | 428 | 428 |
| 1986 | 1 | 10 | 85 | 54 | 206 | 265 | 289 | 339 | 368 | 391 | 382 | 388 | 395 | 395 | 395 | 395 | 395 |
| 1987 | 1 | 10 | 55 | 90 | 143 | 241 | 279 | 299 | 316 | 342 | 343 | 362 | 376 | 376 | 376 | 376 | 376 |
| 1988 | 1 | 15 | 50 | 98 | 135 | 197 | 277 | 315 | 339 | 343 | 359 | 365 | 376 | 376 | 376 | 376 | 376 |
| 1989 | 1 | 15 | 100 | 154 | 175 | 209 | 252 | 305 | 367 | 377 | 359 | 395 | 396 | 396 | 396 | 396 | 396 |
| 1990 | 1 | 8 | 48 | 219 | 198 | 258 | 288 | 309 | 428 | 370 | 403 | 387 | 440 | 440 | 440 | 440 | 440 |
| 1991 | 1 | 11 | 37 | 147 | 210 | 244 | 300 | 324 | 336 | 343 | 382 | 366 | 425 | 425 | 425 | 425 | 425 |
| 1992 | 1 | 7 | 30 | 128 | 224 | 296 | 327 | 355 | 345 | 367 | 341 | 361 | 430 | 470 | 470 | 470 | 450 |
| 1993 | 1 | 8 | 25 | 81 | 201 | 265 | 323 | 354 | 358 | 381 | 369 | 396 | 393 | 374 | 403 | 400 | 400 |
| 1994 | 1 | 10 | 25 | 75 | 151 | 254 | 318 | 371 | 347 | 412 | 382 | 407 | 410 | 410 | 410 | 410 | 410 |
| 1995 | 1 | 18 | 25 | 66 | 138 | 230 | 296 | 346 | 388 | 363 | 409 | 414 | 422 | 410 | 410 | 405 | 447 |
| 1996 | 1 | 18 | 25 | 76 | 118 | 188 | 261 | 316 | 346 | 374 | 390 | 390 | 384 | 398 | 398 | 398 | 398 |
| 1997 | 1 | 18 | 25 | 96 | 118 | 174 | 229 | 286 | 323 | 370 | 378 | 386 | 360 | 393 | 391 | 391 | 391 |
| 1998 | 1 | 18 | 25 | 74 | 147 | 174 | 217 | 242 | 278 | 304 | 310 | 359 | 340 | 344 | 385 | 363 | 375 |
| 1999 | 1 | 18 | 25 | 102 | 150 | 223 | 240 | 264 | 283 | 315 | 345 | 386 | 386 | 386 | 382 | 382 | 407 |
| 2000 | 1 | 18 | 25 | 102 | 150 | 223 | 240 | 264 | 283 | 315 | 345 | 386 | 386 | 386 | 382 | 382 | 407 |
| 2001 | 1 | 18 | 25 | 75 | 178 | 238 | 247 | 296 | 307 | 314 | 328 | 351 | 376 | 406 | 414 | 425 | 425 |
| 2002 | 1 | 10 | 23 | 57 | 177 | 241 | 275 | 302 | 311 | 314 | 328 | 341 | 372 | 405 | 415 | 467 | 409 |
| 2003 | 1 | 10 | 55 | 98 | 159 | 211 | 272 | 305 | 292 | 331 | 337 | 347 | 356 | 381 | 414 | 425 | 441 |




















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Table 3．2．2．5 Norwegian spring spawning herring．Natural mortality used for all years．

Table 3.2.3.1.1 Norwegian Spring-spawning herring. Estimates from the acoustic surveys on the spawning stock in February-March. Numbers in millions.

| Year <br> Age | $\mathbf{1 9 8 8}$ | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 |  | 101 | 183 | 44 |  |  | 16 |  | 407 |  |  | 106 | 1516 |
| 3 | 255 | 5 | 187 | 59 |  |  | 128 | 1792 | 231 |  |  | 1366 | 690 |
| 4 | 146 | 373 | 0 | 54 |  |  | 676 | 7621 | 7638 |  | 381 | 337 | 1996 |
| 5 | 6805 | 103 | 345 | 12 |  |  | 1375 | 3807 | 11243 |  | 1905 | 1286 | 164 |
| 6 | 202 | 5402 | 112 | 354 |  |  | 476 | 2151 | 2586 |  | 10640 | 2979 | 592 |
| 7 |  | 182 | 4489 | 122 |  |  | 63 | 322 | 957 |  | 6708 | 11791 | 1997 |
| 8 |  |  | 146 | 4148 |  |  | 13 | 20 | 471 |  | 1280 | 7534 | 7714 |
| 9 |  |  |  | 102 |  |  | 140 | 1 | 0 |  | 434 | 1912 | 4240 |
| 10 |  |  |  |  |  |  | 35 | 124 | 0 |  | 130 | 568 | 553 |
| 11 |  |  |  |  |  |  | 1820 | 63 | 165 |  | 39 | 132 | 71 |
| 12 |  |  |  |  |  |  |  | 2573 | 0 |  | 0 | 0 | 3 |
| 13 |  |  |  |  |  |  |  |  | 2024 |  | 175 | 0 | 0 |
| 14 |  |  |  |  |  |  |  |  |  |  | 0 | 392 | 6 |
| $15+$ |  |  |  |  |  |  |  |  |  |  | 804 | 437 | 361 |
| Total | 7408 | 6166 | 5462 | 4895 | - | - | 4742 | 18474 | 25756 | - | 22496 | 28840 | 19903 |

In 1992, 1993 and 1997 there was no estimate due to poor weather conditions.
No surveys have been conducted after 2000.
Table 3.2.3.2.1 Norwegian Spring-spawning herring. Estimates obtained on the acoustic surveys in the wintering areas in November-December. Numbers in millions.

| Year <br> Age | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2 *}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 |  |  | 72 |  | 380 |  | 9 | 65 | 74 | 56 | 362 |
| 2 | 36 | 1518 | 16 | 183 | 1465 | 73 | 1207 | 159 | 322 | 522 | 50 |
| 3 | 1247 | 2389 | 3708 | 5133 | 3008 | 661 | 441 | 2425 | 1522 | 3916 | 276 |
| 4 | 1317 | 3287 | 4124 | 5274 | 13180 | 1480 | 1833 | 296 | 5260 | 1528 | 1659 |
| 5 | 173 | 1267 | 2593 | 1839 | 5637 | 6110 | 3869 | 837 | 165 | 2615 | 624 |
| 6 | 16 | 13 | 1096 | 1040 | 994 | 4458 | 12052 | 2066 | 497 | 82 | 1029 |
| 7 | 208 | 13 | 34 | 308 | 552 | 1843 | 8242 | 6601 | 1869 | 338 | 32 |
| 8 | 139 | 158 | 25 | 19 | 92 | 743 | 2068 | 4168 | 4785 | 864 | 188 |
| 9 | 3742 | 26 | 196 | 13 | 0 | 66 | 629 | 755 | 3635 | 3160 | 516 |
| 10 | 69 | 4435 | 29 | 111 | 7 | 0 | 111 | 212 | 668 | 2216 | 1831 |
| 11 |  |  | 3239 | 39 | 41 | 0 | 14 | 0 | 205 | 384 | 911 |
| 12 |  |  |  | 907 | 15 | 126 | 0 | 15 | 0 | 127 | 184 |
| 13 |  |  |  |  | 393 | 0 | 392 | 0 | 0 | 0 | 0 |
| $14+$ |  |  |  |  |  | 842 | 221 | 146 | 168 | 18 | 0 |
| Total | 6947 | 13178 | 15209 | 15246 | 25384 | 16411 | 3144 | 17754 | 19152 | 16132 | 7345 |


| Year <br> Age | 2003** |
| :--- | ---: |
| 1 | 586 |
| 2 | 406 |
| 3 | 2167 |
| 4 | 10670 |
| 5 | 13237 |
| 6 | 1047 |
| 7 | 678 |
| 8 | 41 |
| 9 | 134 |
| 10 | 301 |
| 11 | 1214 |
| 12 | 502 |
| 13 | 10 |
| $14+$ | 37 |
| Total | 31030 |

*Much of the youngest yearclasses ( $-98,-99$ ) wintered outside the fjords this winter and are not included in the estimate
${ }^{* *} 2003$ is the first year with a combined estimate from the Tysfjord, Ofotfjord and oceanic areas off Vesterålen/Troms.

Table 3.2.3.2.2 Norwegian Spring Spawning herring. Estimates obtained on the acoustic surveys in the wintering areas in January. Numbers in millions. No surveys carried out in 2000-2002.

| Year <br> Age | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 90 |  |  |  | 73 |  |  |  | 214 |
| 3 | 220 | 410 | 61 | 642 | 47 | 315 |  | 267 | 1358 |
| 4 | 70 | 820 | 1905 | 3431 | 3781 | 10442 |  | 1938 | 199 |
| 5 | 20 | 260 | 2048 | 4847 | 4013 | 13557 |  | 4162 | 1455 |
| 6 | 180 | 60 | 256 | 1503 | 2445 | 4312 |  | 9647 | 4452 |
| 7 | 150 | 510 | 27 | 102 | 1215 | 1271 |  | 6974 | 12971 |
| 8 | 5500 | 120 | 269 | 29 | 42 | 290 |  | 1518 | 7226 |
| 9 | 440 | 4690 | 182 | 161 | 24 | 22 |  | 743 | 1876 |
| 10 |  | 30 | 5691 | 131 | 267 | 25 |  | 16 | 499 |
| 11 |  |  | 128 | 3679 | 29 | 200 |  | 4 | 16 |
| 12 |  |  |  |  | 4326 | 58 |  | 0 | 16 |
| 13 |  |  |  |  |  | 1146 |  | 181 | 0 |
| 14 |  |  |  |  |  |  |  | 7 | 156 |
| $15+$ |  |  |  |  |  |  |  | 314 | 220 |
| Total | 6670 | 6900 | 10567 | 14598 | 16189 | 31638 | - | 25985 | 30444 |

In 1997 there was no estimate due to poor weather conditions.
Table 3.2.3.3.1 Norwegian spring-spawning herring. Estimates from the international acoustic surveys on the feeding areas in the Norwegian Sea in May. Numbers in millions.

| $\begin{aligned} & \hline \text { Year } \\ & \text { Age } \\ & \hline \end{aligned}$ | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0 | 0 | 24 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 1404 | 215 | 157 | 1540 | 677 |
| 3 | 4114 | 1169 | 367 | 2191 | 1353 | 8312 | 6343 |
| 4 | 22461 | 3599 | 1099 | 322 | 2783 | 1430 | 9619 |
| 5 | 13244 | 18867 | 4410 | 965 | 92 | 1463 | 1418 |
| 6 | 4916 | 13546 | 16378 | 3067 | 384 | 179 | 779 |
| 7 | 2045 | 2473 | 10160 | 11763 | 1302 | 204 | 375 |
| 8 | 424 | 1771 | 2059 | 6077 | 7194 | 3215 | 874 |
| 9 | 14 | 178 | 804 | 853 | 5344 | 5433 | 1941 |
| 10 | 7 | 77 | 183 | 258 | 1689 | 1220 | 2500 |
| 11 | 155 | 288 | 0 | 5 | 271 | 94 | 1423 |
| 12 | 0 | 415 | 0 | 14 | 0 | 178 | 61 |
| 13 | 3134 | 60 | 112 | 0 | 114 | 0 | 78 |
| 14 |  | 2472 | 0 | 158 | 1135 | 0 | 28 |
| 15+ |  |  | 415 | 128 | 1135 | 85 | 26 |
| Total | 50514 | 44915 | 37415 | 26016 | 22953 | 23353 | 26142 |
| Year | 2003 |  |  |  |  |  |  |
| Age |  |  |  |  |  |  |  |
| 1 | 32073 |  |  |  |  |  |  |
| 2 | 8115 |  |  |  |  |  |  |
| 3 | 6561 |  |  |  |  |  |  |
| 4 | 9985 |  |  |  |  |  |  |
| 5 | 9961 |  |  |  |  |  |  |
| 6 | 1499 |  |  |  |  |  |  |
| 7 | 732 |  |  |  |  |  |  |
| 8 | 146 |  |  |  |  |  |  |
| 9 | 228 |  |  |  |  |  |  |
| 10 | 1865 |  |  |  |  |  |  |
| 11 | 2359 |  |  |  |  |  |  |
| 12 | 1769 |  |  |  |  |  |  |
| 13 |  |  |  |  |  |  |  |
| 14 | 287 |  |  |  |  |  |  |
| 15+ | 45 |  |  |  |  |  |  |
| Total | 75625 |  |  |  |  |  |  |

Table 3.2.3.4.1. Norwegian spring-spawning herring. Acoustic estimates (billion individuals) of immature herring in the Barents Sea in May/June. No survey in 2003. 1990-2002.See footnotes.

| Year | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}^{1}$ | $\mathbf{1 9 9 7}^{2}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 24.3 | 32.6 | 102.7 | 6.6 | 0.5 | 0.1 | 2.6 | 9.5 | 49.5 | 105.4 | 0.3 | 0.5 |
| 2 | 5.2 | 14.0 | 25.8 | 59.2 | 7.7 | 0.25 | 0.04 | 4.7 | 4.9 | 27.9 | 7.6 | 3.9 |
| 3 |  | 5.7 | 1.5 | 18.0 | 8.0 | 1.8 | 0.4 | 0.01 | 0.00 | 0.00 | 8.8 | 0.0 |
| 4 |  |  |  | 1.7 | 1.1 | 0.6 | 0.35 | 0.01 | 0.00 | 0.00 | 0.00 | 0.0 |
| 5 |  |  |  |  |  | 0.03 | 0.05 | 0.00 | 0.00 | 0.00 | 0.00 | 0.0 |

${ }^{1}$ Average of Norwegian and Russian estimates
${ }^{2}$ Combination of Norwegian and Russian estimates as described in 1998 WG report, since then only Russian estimates

Table 3.2.3.4.2. Norwegian spring-spawning herring. Acoustic estimates (billion individuals) of immature herring in the Barents Sea in August/September 2000-2001 and 2003 and October/December 2002.

| Year <br> Age | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ |
| :--- | :--- | ---: | ---: | ---: |
| 0 | 0.00 | 0.00 | 106.5 | 0.00 |
| 1 | 14.7 | 0.5 | 1,3 | 99.9 |
| 2 | 11.5 | 10.5 | 0.00 | 4.3 |
| 3 | 0.00 | 1.7 | 0.00 | 2.5 |
| 4 | 0.00 | 0.00 | 0.00 | 0.00 |
| 5 | 0.00 | 0.00 | 0.00 | 0.00 |

Table 3.2.3.4.3 Norwegian spring spawners. Acoustic abundance (TS $=20 \log \mathrm{~L}-71.9$ ) of 0-group herring in Norwegian coastal waters in 1975-2002 (numbers in millions).

| Year | Area |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | South of $62^{\circ} \mathrm{N}$ | 62 | N-65 | N | 65 | N-68 | N | North of 68 | $30^{\prime}$ |  |
| 1975 |  |  | 164 |  |  | 346 |  | 28 |  | 538 |
| 1976 |  |  | 208 |  |  | 1305 |  | 375 |  | 1888 |
| 1977 |  |  | 35 |  |  | 153 |  | 19 |  | 207 |
| 1978 |  |  | 151 |  |  | 256 |  | 196 |  | 603 |
| 1979 |  |  | 455 |  |  | 1130 |  | 144 |  | 1729 |
| 1980 |  |  | 6 |  |  | 2 |  | 109 |  | 117 |
| 1981 |  |  | 132 |  |  | 1 |  | 1 |  | 134 |
| 1982 |  |  | 32 |  |  | 286 |  | 1151 |  | 1469 |
| 1983 |  |  | 162 |  |  | 2276 |  | 4432 |  | 6866 |
| 1984 |  |  | 2 |  |  | 234 |  | 465 |  | 701 |
| 1985 |  |  | 221 |  |  | 177 |  | 104 |  | 502 |
| 1986 |  |  | 5 |  |  | 72 |  | 127 |  | 204 |
| 1987 |  |  | 327 |  |  | 26 |  | 57 |  | 410 |
| 1988 |  |  | 14 |  |  | 552 |  | 708 |  | 1274 |
| 1989 |  |  | 575 |  |  | 263 |  | 2052 |  | 2890 |
| 1990 |  |  | 75 |  |  | 146 |  | 788 |  | 1009 |
| 1991 |  |  | 80 |  |  | 299 |  | 2428 |  | 2807 |
| 1992 |  |  | 73 |  |  | 1993 |  | 621 |  | 2891 |
| 1993 | 290 |  | 109 |  |  | 140 |  | 288 |  | 827 |
| 1994 | 157 |  | 452 |  |  | 323 |  | 6168 |  | 7101 |
| 1995 | 0 |  | 27 |  |  | 2 |  | 0 |  | 29 |
| 1996 | 0 |  | 20 |  |  | 114 |  | 8800 |  | 8934 |
| 1997 | 208 |  | 69 |  |  | 544 |  | 5244 |  | 6065 |
| 1998 | 424 |  | 273 |  |  | 442 |  | 11640 |  | 12779 |
| 1999 | 121 |  | 658 |  |  | 271 |  | 6329 |  | 7379 |
| 2000 | 570 |  | 127 |  |  | 996 |  | 7237 |  | 8930 |
| 2001 | 89 |  | 324 |  |  | 134 |  | 1421 |  | 1968 |
| 2002 | 67 |  | 1227 |  |  | 284 |  | 3573 |  | 5151 |
| 2003 |  |  |  |  |  |  |  |  |  | 27515 |

*A new survey design was introduced in 2003 which resulted in better coverage and higher numbers. The number of 2003 is thus not directly comparable with the earlier years.. A new area system was also introduced and the data is not be given by area in 2003. From 2004 the two series will probably be presented in two tables.

Table 3.2.3.4.4 Norwegian spring-spawning herring. Abundance indices for 0-group herring in the Barents Sea, 1974-2002.

| Year | Log index | Year | Log index |
| ---: | ---: | ---: | ---: |
| 1974 | 0.01 | 1989 | 0.58 |
| 1975 | 0.00 | 1990 | 0.31 |
| 1976 | 0.00 | 1991 | 1.19 |
| 1977 | 0.01 | 1992 | 1.05 |
| 1978 | 0.02 | 1993 | 0.75 |
| 1979 | 0.09 | 1994 | 0.28 |
| 1980 | 0.00 | 1995 | 0.16 |
| 1981 | 0.00 | 1996 | 0.65 |
| 1982 | 0.00 | 1997 | 0.39 |
| 1983 | 1.77 | 1998 | 0.59 |
| 1984 | 0.34 | 1999 | 0.41 |
| 1985 | 0.23 | 2000 | 0.30 |
| 1986 | 0.00 | 2001 | 0.13 |
| 1987 | 0.00 | 2002 | 0.53 |
| 1988 | 0.30 | 2003 | 0.68 |
|  |  |  |  |

Table 3.2.3.5.1 The indices for herring larvae on the Norwegian shelf for the period 1981-2003 ( $\mathrm{N}^{*} 10^{-12}$ )

| Year | Index1 | Index 2 | Year | Index 1 | Index 2 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1981 | 0.3 |  | 1993 | 24.7 | 78.0 |
| 1982 | 0.7 |  | 1994 | 19.5 | 48.6 |
| 1983 | 2.5 |  | 1995 | 18.2 | 36.3 |
| 1984 | 1.4 |  | 1996 | 27.7 | 81.7 |
| 1985 | 2.3 |  | 1997 | 66.6 | 147.5 |
| 1986 | 1.0 |  | 1998 | 42.4 | 138.6 |
| 1987 | 1.3 | 4.0 | 1999 | 19.9 | 73.0 |
| 1988 | 9.2 | 25.5 | 2000 | 19.8 | 127.5 |
| 1989 | 13.4 | 28.7 | 2001 | 40.7 | 131.9 |
| 1990 | 18.3 | 29.2 | 2002 | 27.1 | 113.9 |
| 1991 | 8.6 | 23.5 | 2003 | 3.7 | 18.9 |
| 1992 | 6.3 | 27.8 | 2004 | 56.4 | 175.7 |
|  |  |  |  |  |  |

Index 1. The total number of herring larvae found during the cruise.
Index 2. Back-calculated number of newly hatched larvae with $10 \%$ daily moratlity. The larval age is estimated from the duration of the yolksac stages and the size of the larvae.

Table 3.2.4.1 Tagging data used in the SeaStar runs. Please note that the tagging data for 2002 was considered an outlier and thus not included in the analysis - Norwegian spring spawning herring.


Tagging data for the 1984 year class


Table 3.2.4.1 continued. Tagging data for the 1985 year class

*1985+ group
Tagging data for the 1986 year class


Table 3.2.4.1 continued
Tagging data for the 1987 year class

*1987+group
Tagging data for the 1988 year class


Table 3.2.4.1 continued Tagging data for the 1989 year class

| $\begin{aligned} & \text { ঝ } \\ & \underset{\sim}{\circ} \end{aligned}$ | z | $\begin{aligned} & \text { Z } \\ & \hline \end{aligned}$ | Recovered |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \pi \\ & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ |  |  | $\begin{array}{\|c\|} \hline 2000 \\ \text { release } \end{array}$ | $\begin{array}{c\|} 1998 \\ \text { release } \end{array}$ | $\begin{array}{c\|} 1997 \\ \text { release } \end{array}$ | $\begin{gathered} 1996 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1995 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1994 \\ \text { release } \end{gathered}$ | $\left\lvert\, \begin{gathered} 1993 \\ \text { release } \end{gathered}\right.$ |
| 1993 | 7584 |  |  |  |  |  |  |  |  |
| 1994 | 11873 |  |  |  |  |  |  |  |  |
| 1995 | 2348 | 9463 |  |  |  |  |  |  | 4 |
| 1996 | 5170 | 4636 |  |  |  |  |  | 5 | 1 |
| 1997 | 4103 | 3346 |  |  |  |  | 0 | 7 | 2 |
| 1998 | 1176 | 1183 |  |  |  | 1 | 0 | 0 | 0 |
| 1999 |  | 1179 |  |  | 1 | 1 | 0 | 0 | 1 |
| 2000 | 470 | 790 |  | 1 | 0 | 0 | 0 | 2 | 0 |
| 2001 |  | 841 |  | 0 | 0 | 2 | 0 | 1 | 1 |
| 2002 | 319 | 286 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 59 | 460 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |

Tagging data for the 1990 year class

| $\begin{aligned} & \widehat{\aleph} \\ & \underset{\sim}{\circ} \end{aligned}$ | $\begin{aligned} & \text { Z } \\ & \text { Z } \\ & \text { B } \end{aligned}$ | Z Z 0 0 | Recovered |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \widetilde{0} \\ & \text { ¿} \\ & \text { O} \end{aligned}$ |  |  | $\begin{array}{\|c\|} 2000 \\ \text { release } \end{array}$ | $\begin{gathered} 1998 \\ \text { release } \end{gathered}$ | $\begin{array}{c\|} 1997 \\ \text { release } \end{array}$ | $\begin{array}{\|c} 1996 \\ \text { release } \end{array}$ | $\begin{array}{\|c\|c\|} 1995 \\ \text { release } \end{array}$ | $\begin{gathered} 1994 \\ \text { release } \end{gathered}$ | $\begin{array}{\|c} 1993 \\ \text { release } \end{array}$ |
| 1994 | 10784 |  |  |  |  |  |  |  |  |
| 1995 | 3868 |  |  |  |  |  | 0 | 3 |  |
| 1996 | 6171 | 9009 |  |  |  | 3 | 3 | 9 |  |
| 1997 | 4057 | 9830 |  |  | 2 | 3 | 3 | 7 |  |
| 1998 | 2381 | 2828 |  | 2 | 3 |  | 1 | 1 |  |
| 1999 |  | 3402 |  | 3 | 1 | 2 | 2 | 1 |  |
| 2000 | 1219 | 3146 |  | 1 | 0 | 2 | 2 | 0 |  |
| 2001 |  | 1052 |  | 0 | 0 | 0 | 0 | 2 |  |
| 2002 | 1605 | 1348 | 0 | 1 | 0 | 1 | 0 | 0 |  |
| 2003 | 56 | 1129 | 0 | 0 | 0 | 1 | 0 | 1 |  |

Tagging data for the 1991 year class


Table 3.2.4.1 continued
Tagging data for the 1992 year class


| $\underset{\cong}{\aleph}$ | 艺 | $\begin{aligned} & \text { Z } \\ & \text { Z } \\ & \text { B } \end{aligned}$ | Recovered |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { §̀ } \\ & \text { O} \\ & \text { ¢ } \end{aligned}$ |  |  | $\begin{gathered} 2000 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1998 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1997 \\ \text { release } \end{gathered}$ |  |
| 1997 1998 1999 2000 2001 2002 | $\begin{array}{r} 976 \\ 2015 \\ 2673 \\ \\ 2832 \\ \hline \end{array}$ | $\begin{aligned} & 8046 \\ & 9049 \\ & 3994 \\ & 5577 \end{aligned}$ |  |  |  | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 5 \end{aligned}$ |
| 2003 | 1020 | 6612 |  |  |  | 1 |

Tagging data for the 1994 year class

| $\begin{aligned} & \text { § } \\ & \text { ® } \end{aligned}$ | $\stackrel{Z}{Z}$ | $\stackrel{Z}{Z}$ | Recovered |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \ddot{0} \\ & \hat{0} \\ & 0 \\ & \varnothing \ddot{Q} \end{aligned}$ | $\begin{gathered} \stackrel{\rightharpoonup}{0} \\ 0 \\ 00 \\ 0 \\ 0 \\ 0 \end{gathered}$ |  | $\begin{gathered} 2000 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1998 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1997 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1996 \\ \text { release } \end{gathered}$ |
| $\begin{aligned} & 1998 \\ & 1999 \\ & 2000 \\ & 2001 \\ & 2002 \end{aligned}$ | $\begin{aligned} & 3752 \\ & 2278 \end{aligned}$ | $\begin{aligned} & 2450 \\ & 1104 \end{aligned}$ | 12 |  |  |  |
| 2003 | 442 | 2154 |  |  |  |  |

Tagging data for the 1995 year class

| $\underset{\cong}{\S}$ | Z | $\underset{y}{Z}$ | Recovered |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $\begin{gathered} 2000 \\ \text { release } \end{gathered}$ | $\begin{gathered} 1998 \\ \text { release } \end{gathered}$ | $\begin{aligned} & 1997 \\ & \text { release } \end{aligned}$ | $\begin{gathered} 1996 \\ \text { release } \end{gathered}$ |
| $\begin{aligned} & 2000 \\ & 2001 \\ & 2002 \end{aligned}$ | 505 197 | $\begin{aligned} & 276 \\ & 250 \end{aligned}$ |  |  |  |  |
| 2003 | 263 | 747 | 2 |  |  |  |

Table 3.3.2.1 Summary of SeaStar exploratory runs. s1-s4 means survey 1 to survey 6, log-lik means log-likelihood value per term.

Default NoTags EstimateM Oldest10 CorrAgeBias SurvSel NoLarvae

| Spawning | stock | 6.966 | 8.623 | 5.828 | 10.245 | 7.741 | 6.488 | 5.872 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| cat1 |  | 0.837 | 0.727 | 0.982 | 0.592 | 0.750 | 0.782 | 0.899 |
| cat2 |  | 0.786 | 0.667 | 0.940 | 0.592 | 0.703 | 0.740 | 0.877 |
| cat3 |  | 0.867 | 0.776 | 1.023 | 0.685 | 0.791 | 0.856 | 0.917 |
| cat5 |  | 1.013 | 0.849 | 1.208 | 0.705 | 0.929 | 1.029 | 1.122 |
| cat4 |  | 0.392 | 0.344 | 1.035 | 0.315 | 0.370 | 0.360 | 0.426 |
| cat6 |  | 0.492 | 0.427 | 0.969 | 0.380 | 0.468 | 0.450 | 0.553 |
| catZero |  | 0.003 | 0.003 | 0.032 | 0.002 | 0.003 | 0.003 | 0.003 |
| Log-lik, s 1 | -1.606 | -1.573 | -1.603 | -1.587 | -1.567 | -1.109 | -1.635 |  |
| Log-lik, s 2 | -1.344 | -1.371 | -1.344 | -1.579 | -1.358 | -0.959 | -1.350 |  |
| Log-lik, s 3 | -1.768 | -1.770 | -1.769 | -1.769 | -1.739 | -1.402 | -1.779 |  |
| Log-lik, s 4 | -3.247 | -3.272 | -3.274 | -3.271 | -3.271 | -3.247 | -3.250 |  |
| Log-lik, s 5 | -1.817 | -1.795 | -1.789 | -1.759 | -1.853 | -1.418 | -1.798 |  |
| Log-lik, s 6 | -3.355 | -3.386 | -3.480 | -3.382 | -3.385 | -3.405 | -3.366 |  |
| No data, s 1 | 20 | 20 | 20 | 16 | 20 | 20 | 20 |  |
| No data, s 2 | 50 | 50 | 50 | 39 | 50 | 50 | 50 |  |
| No data, s 3 | 18 | 18 | 18 | 14 | 18 | 18 | 18 |  |
| No data, s | 4 | 18 | 18 | 18 | 18 | 18 | 18 | 18 |
| No data, s 5 | 48 | 48 | 48 | 40 | 48 | 48 | 48 |  |
| No data, s 6 | 9 | 9 | 9 | 9 | 9 | 9 | 9 |  |
| M |  | 0.150 | 0.150 | 0.115 | 0.150 | 0.150 | 0.150 | 0.150 |
| MYoung |  | 0.900 | 0.900 | 0.188 | 0.900 | 0.900 | 0.900 | 0.900 |

Table XXX. Summary of SeaStar exploratory runs. s1-s4 means survey 1 to survey 6 , log-lik means loglikelihood value per term.

Default NoTags EstimateM Oldest10 CorrAgeBias SurvSel NoLarvae

| Spawning | stock | 6.966 | 8.623 | 5.828 | 10.245 | 7.741 | 6.488 | 5.872 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| cat1 |  | 0.837 | 0.727 | 0.982 | 0.592 | 0.750 | 0.782 | 0.899 |
| cat2 |  | 0.786 | 0.667 | 0.940 | 0.592 | 0.703 | 0.740 | 0.877 |
| cat 3 |  | 0.867 | 0.776 | 1.023 | 0.685 | 0.791 | 0.856 | 0.917 |
| cat5 |  | 1.013 | 0.849 | 1.208 | 0.705 | 0.929 | 1.029 | 1.122 |
| cat 4 |  | 0.392 | 0.344 | 1.035 | 0.315 | 0.370 | 0.360 | 0.426 |
| cat 6 |  | 0.492 | 0.427 | 0.969 | 0.380 | 0.468 | 0.450 | 0.553 |
| catZero |  | 0.003 | 0.003 | 0.032 | 0.002 | 0.003 | 0.003 | 0.003 |
| Log-lik, | S 1 | -1.606 | -1.573 | -1.603 | -1.587 | -1.567 | -1.109 | -1.635 |
| Log-lik, | s 2 | -1.344 | -1.371 | -1.344 | -1.579 | -1.358 | -0.959 | -1.350 |
| Log-lik, | s 3 | -1.768 | -1.770 | -1.769 | -1.769 | -1.739 | -1.402 | -1.779 |
| Log-lik, | S 4 | -3.247 | $-3.272$ | -3.274 | -3.271 | -3.271 | -3.247 | -3.250 |
| Log-lik, | s 5 | -1.817 | -1.795 | -1.789 | -1.759 | -1.853 | -1.418 | -1.798 |
| Log-lik, | s 6 | $-3.355$ | $-3.386$ | -3.480 | $-3.382$ | $-3.385$ | -3.405 | -3.366 |
| No data, | s 1 | 20 | 20 | 20 | 16 | 20 | 20 | 20 |
| No data, | s 2 | 50 | 50 | 50 | 39 | 50 | 50 | 50 |
| No data, | s 3 | 18 | 18 | 18 | 14 | 18 | 18 | 18 |
| No data, | S 4 | 18 | 18 | 18 | 18 | 18 | 18 | 18 |


| No data, s | 5 | 48 | 48 | 48 | 40 | 48 | 48 | 48 |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| No data, s | 6 | 9 | 9 | 9 | 9 | 9 | 9 | 9 |
| M |  | 0.150 | 0.150 | 0.115 | 0.150 | 0.150 | 0.150 | 0.150 |
| MYoung |  | 0.900 | 0.900 | 0.188 | 0.900 | 0.900 | 0.900 | 0.900 |

Table 3.4.1.1. Model settings used in final ISVPA run for NSS herring - Norwegian spring spawning herring.

| Settings/options for the ISVPA run | WG 2003 | WG 2004 |
| :--- | :--- | :--- |
| Numbers of age structured tuning series | 6 | 6 |
| Version of the model | Catch-controlled | Catch-controlled |
| Numbers of biomass tuning series | 0 | 0 |
| Constraint | Unbiased model <br> description of <br> logarithmic <br> catch-at-age | Unbiased model <br> description of <br> logarithmic <br> catch-at-age |
| Number of years with a separable constraint | $1986-2002$ | $1950-2003$ |
| Reference age for separable constraint | no | no |
| Constant selection pattern | Equal to that for <br> previous age | Equal to that for <br> previous age |
| Selection patterns |  |  |
| Sor two perisds |  |  |$|$| $5-14$ |
| :--- |
| Age span for calculation of reference F |

Table 3.4.1.2 NSS herring. Estimates of stock abundance by ISVPA - Norwegian spring spawning herring.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 30138411 | 16104151 | 11500353 | 3920266 | 4495539 | 7722200 | 6835114 | 1624898 | 2376953 | 2946662 | 2563065 | 6147022 | 6703771 | 1196992 | 1832064 | 4063072 |
| 1951 | 321818228 | 10978107 | 6164882 | 9642203 | 3202757 | 3697250 | 6139084 | 5299858 | 1324807 | 1963664 | 2434628 | 2125430 | 5110344 | 5428301 | 968658 | 6234970 |
| 1952 | 59433332 | 125990650 | 4208314 | 5300040 | 7943053 | 2596696 | 3029732 | 4805614 | 4003128 | 1068743 | 1613417 | 1999853 | 1729549 | 4163331 | 4349327 | 5228356 |
| 1953 | 30542849 | 18329683 | 50437844 | 3585669 | 4505658 | 6277869 | 2108545 | 2417991 | 3783502 | 3094930 | 846399 | 1309173 | 1621372 | 1389554 | 3410388 | 8148533 |
| 1954 | 31362621 | 9194585 | 7081640 | 42725632 | 3042981 | 3784447 | 5073507 | 1738860 | 1978297 | 2965086 | 2297465 | 671260 | 1042206 | 1308227 | 1104339 | 5890969 |
| 1955 | 10298732 | 8242364 | 3192812 | 5848166 | 35442516 | 2486544 | 3038356 | 3911935 | 1377806 | 1517373 | 2143495 | 1550035 | 495746 | 803704 | 1002611 | 4616271 |
| 1956 | 6854115 | 2356458 | 3025841 | 2661799 | 4777135 | 28608328 | 2034147 | 2439237 | 3112183 | 1106752 | 1126589 | 1570682 | 1145610 | 372234 | 613267 | 3285015 |
| 1957 | 8726573 | 1496307 | 558208 | 2496283 | 2057611 | 3820221 | 22252939 | 1648755 | 1910304 | 2433571 | 831335 | 785693 | 1098810 | 834535 | 261937 | 1925278 |
| 1958 | 7137281 | 1449653 | 468394 | 458837 | 1802244 | 1628315 | 3076105 | 17311433 | 1352299 | 1526112 | 1925280 | 633524 | 563810 | 807243 | 596200 | 1599604 |
| 1959 | 3230913 | 1117655 | 164470 | 386915 | 378318 | 1448319 | 1318657 | 2467275 | 13996930 | 1098343 | 1199424 | 1470721 | 453711 | 413468 | 629023 | 1112603 |
| 1960 | 156311627 | 1187003 | 246857 | 127551 | 308157 | 301593 | 1110573 | 1028473 | 1900295 | 11023226 | 863154 | 917036 | 1082168 | 308407 | 284068 | 732438 |
| 1961 | 72099785 | 54892071 | 232330 | 99565 | 92900 | 239164 | 236947 | 866630 | 817212 | 1446433 | 8408813 | 663880 | 668972 | 789022 | 212845 | 376648 |
| 1962 | 26987917 | 19063330 | 20478021 | 171023 | 78182 | 76156 | 191934 | 185944 | 688766 | 657736 | 1118690 | 6562042 | 525298 | 534041 | 620669 | 820305 |
| 1963 | 5382250 | 8370244 | 7086609 | 15915022 | 139779 | 64416 | 58868 | 146459 | 149003 | 537997 | 517319 | 854320 | 4893283 | 411122 | 408906 | 845568 |
| 1964 | 65619925 | 836998 | 2098946 | 5394045 | 12922779 | 115392 | 53773 | 47328 | 109080 | 119620 | 363140 | 359444 | 573799 | 3354731 | 280008 | 1779087 |
| 1965 | 35893563 | 24939430 | 199828 | 1700261 | 4272528 | 9224761 | 86608 | 44891 | 37953 | 70786 | 75775 | 223866 | 232930 | 351929 | 2170484 | 552178 |
| 1966 | 1992966 | 12182425 | 8320080 | 88590 | 1225740 | 3147565 | 5899068 | 56454 | 24815 | 25801 | 43206 | 28110 | 99445 | 100474 | 174230 | 828809 |
| 1967 | 18411013 | 387660 | 3883065 | 5260491 | 51294 | 622119 | 1497501 | 2401299 | 13429 | 8092 | 6064 | 12881 | 13990 | 21486 | 19589 | 123423 |
| 1968 | 1374507 | 1187442 | 112722 | 2050487 | 1508869 | 19471 | 144604 | 238705 | 470356 | 3301 | 1676 | 1972 | 3201 | 3784 | 2257 | 19063 |
| 1969 | 972036 | 280189 | 235187 | 5081 | 20249 | 11544 | 3585 | 37254 | 81045 | 84675 | 986 | 422 | 955 | 435 | 845 | 4224 |
| 1970 | 4348950 | 71859 | 23437 | 27826 | 3631 | 9265 | 5576 | 2436 | 21210 | 38584 | 39482 | 570 | 178 | 637 | 189 | 3596 |
| 1971 | 329396 | 1430591 | 8046 | 14328 | 6694 | 2569 | 4912 | 1738 | 1169 | 5824 | 8903 | 7913 | 0 | 60 | 363 | 0 |
| 1972 | 106156 | 106568 | 527373 | 5237 | 11386 | 4611 | 1877 | 3198 | 447 | 0 | 922 | 1252 | 1754 | 0 | 0 | 0 |
| 1973 | 134959 | 17017 | 30320 | 421094 | 1283 | 6476 | 1667 | 972 | 0 | 0 | 0 | 0 | 0 | 960 | 0 | 0 |
| 1974 | 5181226 | 52639 | 5835 | 23880 | 339060 | 500 | 4176 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 3484678 | 2101556 | 18915 | 4929 | 20330 | 269097 | 192 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 1181702 | 1414469 | 853281 | 13248 | 4120 | 16654 | 203163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 3438676 | 478914 | 574315 | 712858 | 6359 | 3546 | 14334 | 162724 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 2137845 | 1394108 | 192735 | 473812 | 591672 | 5162 | 3052 | 11949 | 130069 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 2549826 | 867653 | 566037 | 163088 | 396528 | 490410 | 3636 | 2627 | 9709 | 107288 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 4839758 | 1034259 | 351550 | 481299 | 138640 | 334926 | 411694 | 2827 | 2261 | 8357 | 89993 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 617367 | 1967189 | 420243 | 296638 | 408864 | 117215 | 280699 | 339655 | 2024 | 1939 | 7193 | 74964 | 0 | 0 | 0 | 0 |
| 1982 | 465161 | 250301 | 792212 | 357842 | 251059 | 343938 | 98847 | 237414 | 284662 | 1422 | 1573 | 6085 | 63627 | 0 | 0 | 0 |
| 1983 | 907877 | 188419 | 101637 | 669044 | 300675 | 211907 | 290224 | 83260 | 199635 | 239401 | 1112 | 1320 | 5203 | 54652 | 0 | 0 |
| 1984 | 127343366 | 366131 | 75541 | 84527 | 556192 | 249961 | 176656 | 243468 | 70463 | 167562 | 199255 | 824 | 0 | . | 46242 | 0 |
| 1985 | 4898450 | 51772865 | 147270 | 60859 | 67754 | 421623 | 198256 | 140324 | 195075 | 53954 | 129065 | 165490 | 709 | 0 | 0 | 15890 |
| 1986 | 6735948 | 1983176 | 20917147 | 106810 | 38002 | 43009 | 242287 | 115904 | 69752 | 109455 | 37162 | 82327 | 96052 | 611 | 0 | 2398 |
| 1987 | 3815091 | 2737752 | 804329 | 17502773 | 75610 | 19256 | 22638 | 111126 | 30179 | 21071 | 22772 | 13923 | 9628 | 8453 | 526 | 0 |
| 1988 | 3616859 | 1547064 | 1090279 | 673945 | 14599612 | 47755 | 13325 | 12937 | 69670 | 14842 | 9323 | 15425 | 4716 | 2257 | 781 | 0 |
| 1989 | 10739134 | 1468726 | 623181 | 880035 | 556822 | 12055403 | 32334 | 8056 | 5602 | 46436 | 4544 | 5410 | 10163 | 1571 | 0 | 0 |
| 1990 | 28037640 | 4364976 | 581071 | 533695 | 754092 | 474019 | 10075324 | 24612 | 6192 | 4192 | 36909 | 2635 | 4026 | 8450 | 1111 | 2025 |
| 1991 | 50065878 | 11398999 | 1764758 | 482846 | 456890 | 638036 | 397922 | 8461982 | 19988 | 3920 | 1719 | 29527 | 1669 | 3299 | 6725 | 127 |
| 1992 | 137219228 | 20353118 | 4632364 | 1511113 | 413010 | 391941 | 535527 | 334269 | 7080258 | 14885 | 2946 | 1399 | 24774 | 0 | 0 | 0 |
| 1993 | 182796101 | 55789079 | 8274106 | 3975436 | 1269919 | 350861 | 336240 | 449817 | 282375 | 5884664 | 10508 | 1943 | 975 | 20177 | 0 | 0 |
| 1994 | 49090778 | 74319266 | 22677530 | 7095234 | 3322545 | 1012066 | 293987 | 286020 | 359697 | 225758 | 4684500 | 9044 | 1672 | 839 | 17366 | 0 |
| 1995 | 18243865 | 19958808 | 30210794 | 19488580 | 6004789 | 2522117 | 718201 | 238583 | 238628 | 274962 | 161228 | 3433212 | 5159 | 1013 | 0 | 1823 |
| 1996 | 5405047 | 7417402 | 8113925 | 25949242 | 16452550 | 4590562 | 1579054 | 403769 | 190961 | 190684 | 171951 | 61081 | 2109002 | 665 | 0 | 0 |
| 1997 | 38016920 | 2197528 | 2996472 | 6951843 | 21672664 | 12703356 | 3078516 | 982181 | 251589 | 159092 | 157286 | 86685 | 36273 | 1039131 | 572 | 0 |
| 1998 | 33342379 | 15456526 | 879535 | 2458064 | 5732135 | 16987811 | 9084312 | 1943496 | 542472 | 160073 | 118358 | 105318 | 46740 | 13482 | 494709 | 0 |
| 1999 | 154995219 | 13556000 | 6231301 | 691781 | 1890822 | 4591997 | 12988420 | 6646504 | 1318865 | 346330 | 98345 | 78360 | 87422 | 26857 | 0 | 458838 |
| 2000 | 113165751 | 63016354 | 5508252 | 5235649 | 562190 | 1502374 | 3553965 | 9690247 | 4640561 | 864819 | 199744 | 71172 | 30298 | 68563 | 13888 | 197233 |
| 2001 | 58193004 | 46009761 | 25611359 | 4663051 | 3986477 | 451472 | 1190387 | 2683691 | 7135099 | 3024675 | 543055 | 105505 | 46173 | 5017 | 0 | 116832 |
| 2002 | 15771098 | 23659510 | 18704849 | 21948999 | 3864457 | 3035211 | 352637 | 935520 | 2034835 | 5362734 | 2132898 | 399062 | 68801 | 36490 | 1205 | 39208 |
| 2003 | 128782619 | 6412050 | 9579823 | 15915644 | 18297624 | 3090686 | 2312963 | 275899 | 718906 | 1507319 | 4005044 | 1523231 | 294762 | 47774 | 24971 | 0 |

Table 3.4.1.3 NSS herring. Estimates of fishing mortality by ISVPA - Norwegian spring spawning herring.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0.0284 | 0.0333 | 0.0376 | 0.0437 | 0.0510 | 0.0627 | 0.0576 | 0.0635 | 0.0604 | 0.0747 | 0.0736 | 0.0647 | 0.0585 | 0.0650 | 0.0650 | 0.0650 |
| 1951 | 0.0237 | 0.0278 | 0.0314 | 0.0365 | 0.0425 | 0.0523 | 0.0480 | 0.0529 | 0.0503 | 0.0622 | 0.0613 | 0.0540 | 0.0488 | 0.0542 | 0.0542 | 0.0542 |
| 1952 | 0.0286 | 0.0336 | 0.0380 | 0.0441 | 0.0515 | 0.0633 | 0.0581 | 0.0641 | 0.0609 | 0.0754 | 0.0743 | 0.0653 | 0.0591 | 0.0656 | 0.0656 | 0.0656 |
| 1953 | 0.0307 | 0.0361 | 0.0408 | 0.0474 | 0.0553 | 0.0680 | 0.0625 | 0.0689 | 0.0655 | 0.0811 | 0.0799 | 0.0702 | 0.0635 | 0.0705 | 0.0705 | 0.0705 |
| 1954 | 0.0566 | 0.0665 | 0.0753 | 0.0878 | 0.1028 | 0.1272 | 0.1166 | 0.1289 | 0.1224 | 0.1526 | 0.1502 | 0.1315 | 0.1185 | 0.1321 | 0.1321 | 0.1321 |
| 1955 | 0.0507 | 0.0597 | 0.0675 | 0.0787 | 0.0921 | 0.1137 | 0.1043 | 0.1152 | 0.1094 | 0.1363 | 0.1341 | 0.1176 | 0.1060 | 0.1181 | 0.1181 | 0.1181 |
| 1956 | 0.0719 | 0.0848 | 0.0961 | 0.1122 | 0.1317 | 0.1635 | 0.1496 | 0.1656 | 0.1571 | 0.1969 | 0.1937 | 0.1691 | 0.1521 | 0.1699 | 0.1699 | 0.1699 |
| 1957 | 0.0636 | 0.0749 | 0.0848 | 0.0990 | 0.1160 | 0.1437 | 0.1316 | 0.1456 | 0.1382 | 0.1727 | 0.1699 | 0.1486 | 0.1338 | 0.1493 | 0.1493 | 0.1493 |
| 1958 | 0.0574 | 0.0675 | 0.0764 | 0.0891 | 0.1043 | 0.1291 | 0.1183 | 0.1308 | 0.1242 | 0.1549 | 0.1524 | 0.1335 | 0.1202 | 0.1341 | 0.1341 | 0.1341 |
| 1959 | 0.0671 | 0.0790 | 0.0895 | 0.1045 | 0.1225 | 0.1519 | 0.1390 | 0.1539 | 0.1460 | 0.1827 | 0.1798 | 0.1571 | 0.1414 | 0.1578 | 0.1578 | 0.1578 |
| 1960 | 0.0829 | 0.0978 | 0.1109 | 0.1298 | 0.1525 | 0.1898 | 0.1735 | 0.1923 | 0.1823 | 0.2292 | 0.2254 | 0.1964 | 0.1764 | 0.1974 | 0.1974 | 0.1974 |
| 1961 | 0.0501 | 0.0589 | 0.0666 | 0.0776 | 0.0908 | 0.1121 | 0.1028 | 0.1136 | 0.1079 | 0.1343 | 0.1322 | 0.1159 | 0.1045 | 0.1164 | 0.1164 | 0.1164 |
| 1962 | 0.0519 | 0.0611 | 0.0691 | 0.0806 | 0.0943 | 0.1165 | 0.1068 | 0.1180 | 0.1121 | 0.1396 | 0.1374 | 0.1204 | 0.1086 | 0.1210 | 0.1210 | 0.1210 |
| 1963 | 0.0734 | 0.0864 | 0.0980 | 0.1145 | 0.1344 | 0.1669 | 0.1526 | 0.1690 | 0.1604 | 0.2010 | 0.1978 | 0.1726 | 0.1552 | 0.1734 | 0.1734 | 0.1734 |
| 1964 | 0.0816 | 0.0962 | 0.1091 | 0.1276 | 0.1499 | 0.1865 | 0.1705 | 0.1890 | 0.1792 | 0.2252 | 0.2215 | 0.1931 | 0.1734 | 0.1940 | 0.1940 | 0.1940 |
| 1965 | 0.1688 | 0.2007 | 0.2294 | 0.2712 | 0.3230 | 0.4116 | 0.3723 | 0.4178 | 0.3935 | 0.5106 | 0.5009 | 0.4280 | 0.3793 | 0.4302 | 0.4302 | 0.4302 |
| 1966 | 0.3089 | 0.3727 | 0.4319 | 0.5217 | 0.6394 | 0.8614 | 0.7591 | 0.8780 | 0.8135 | 1.1526 | 1.1214 | 0.9057 | 0.7769 | 0.9119 | 0.9119 | 0.9119 |
| 1967 | 0.4301 | 0.5263 | 0.6186 | 0.7649 | 0.9711 | 1.4249 | 1.2027 | 1.4637 | 1.3176 | 2.3016 | 2.1780 | 1.5306 | 1.2396 | 1.5458 | 1.5458 | 1.5458 |
| 1968 | 0.4810 | 0.5924 | 0.7010 | 0.8769 | 1.1352 | 1.7693 | 1.4447 | 1.8298 | 1.6084 | 4.0713 | 3.4594 | 1.9372 | 1.4964 | 1.9624 | 1.9624 | 1.9624 |
| 1969 | 0.2571 | 0.3084 | 0.3555 | 0.4257 | 0.5157 | 0.6783 | 0.6046 | 0.6901 | 0.6441 | 0.8761 | 0.8558 | 0.7096 | 0.6176 | 0.7140 | 0.7140 | 0.7140 |
| 1970 | 0.2917 | 0.3512 | 0.4063 | 0.4892 | 0.5972 | 0.7977 | 0.7059 | 0.8125 | 0.7549 | 1.0535 | 1.0265 | 0.8372 | 0.7220 | 0.8427 | 0.8427 | 0.8427 |
| 1971 | 0.1725 | 0.2052 | 0.2346 | 0.2776 | 0.3308 | 0.4221 | 0.3815 | 0.4284 | 0.4034 | 0.5242 | 0.5142 | 0.4389 | 0.0000 | 0.4412 | 0.4412 | 0.0000 |
| 1972 | 0.2584 | 0.3101 | 0.3575 | 0.4282 | 0.5189 | 0.6829 | 0.6085 | 0.6947 | 0.6483 | 0.0000 | 0.8621 | 0.7145 | 0.6216 | 0.0000 | 0.0000 | 0.0000 |
| 1973 | 0.0941 | 0.1111 | 0.1262 | 0.1478 | 0.1739 | 0.2171 | 0.1981 | 0.2200 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.2258 | 0.0000 | 0.0000 |
| 1974 | 0.0256 | 0.0300 | 0.0339 | 0.0394 | 0.0460 | 0.0565 | 0.0519 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1975 | 0.0141 | 0.0166 | 0.0187 | 0.0217 | 0.0252 | 0.0309 | 0.0285 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1976 | 0.0153 | 0.0179 | 0.0203 | 0.0235 | 0.0000 | 0.0000 | 0.0309 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1977 | 0.0163 | 0.0192 | 0.0216 | 0.0251 | 0.0292 | 0.0000 | 0.0330 | 0.0363 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1978 | 0.0120 | 0.0141 | 0.0159 | 0.0184 | 0.0215 | 0.0263 | 0.0000 | 0.0266 | 0.0253 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1979 | 0.0086 | 0.0101 | 0.0114 | 0.0132 | 0.0153 | 0.0188 | 0.0173 | 0.0000 | 0.0000 | 0.0223 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1980 | 0.0062 | 0.0072 | 0.0081 | 0.0094 | 0.0110 | 0.0134 | 0.0124 | 0.0136 | 0.0129 | 0.0000 | 0.0157 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1981 | 0.0107 | 0.0125 | 0.0141 | 0.0163 | 0.0190 | 0.0233 | 0.0214 | 0.0236 | 0.0224 | 0.0277 | 0.0272 | 0.0240 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1982 | 0.0071 | 0.0083 | 0.0094 | 0.0109 | 0.0127 | 0.0155 | 0.0143 | 0.0157 | 0.0149 | 0.0184 | 0.0181 | 0.0160 | 0.0145 | 0.0000 | 0.0000 | 0.0000 |
| 1983 | 0.0147 | 0.0172 | 0.0194 | 0.0225 | 0.0263 | 0.0322 | 0.0296 | 0.0326 | 0.0310 | 0.0382 | 0.0377 | 0.0332 | 0.0301 | 0.0334 | 0.0000 | 0.0000 |
| 1984 | 0.0172 | 0.0202 | 0.0228 | 0.0265 | 0.0308 | 0.0378 | 0.0348 | 0.0383 | 0.0365 | 0.0450 | 0.0443 | 0.0000 | 0.0000 | 0.0000 | 0.0392 | 0.0000 |
| 1985 | 0.0829 | 0.0978 | 0.1109 | 0.1297 | 0.1525 | 0.1898 | 0.1734 | 0.1923 | 0.1823 | 0.2292 | 0.2254 | 0.1964 | 0.0000 | 0.0000 | 0.0000 | 0.1973 |
| 1986 | 0.0002 | 0.0060 | 0.0463 | 0.1270 | 0.2726 | 0.4185 | 0.5705 | 0.7043 | 0.9854 | 1.9502 | 1.8338 | 1.9128 | 2.0377 | 0.0000 | 0.0000 | 2.2033 |
| 1987 | 0.0002 | 0.0050 | 0.0384 | 0.1044 | 0.2211 | 0.3345 | 0.4484 | 0.5451 | 0.7359 | 1.2482 | 1.1986 | 1.2326 | 1.2833 | 2.5641 | 2.5641 | 0.0000 |
| 1988 | 0.0002 | 0.0040 | 0.0308 | 0.0834 | 0.1745 | 0.2607 | 0.3449 | 0.4143 | 0.5456 | 0.8569 | 0.8296 | 0.8484 | 0.8760 | 1.3668 | 1.3668 | 0.0000 |
| 1989 | 0.0000 | 0.0011 | 0.0082 | 0.0217 | 0.0438 | 0.0634 | 0.0813 | 0.0952 | 0.1195 | 0.1674 | 0.1637 | 0.1663 | 0.1699 | 0.2226 | 0.0000 | 0.0000 |
| 1990 | 0.0000 | 0.0009 | 0.0068 | 0.0179 | 0.0361 | 0.0522 | 0.0668 | 0.0781 | 0.0977 | 0.1363 | 0.1334 | 0.1354 | 0.1384 | 0.1803 | 0.1803 | 0.1803 |
| 1991 | 0.0000 | 0.0005 | 0.0035 | 0.0092 | 0.0184 | 0.0265 | 0.0338 | 0.0394 | 0.0491 | 0.0679 | 0.0664 | 0.0674 | 0.0688 | 0.0888 | 0.0888 | 0.0888 |
| 1992 | 0.0000 | 0.0003 | 0.0026 | 0.0069 | 0.0139 | 0.0199 | 0.0254 | 0.0296 | 0.0369 | 0.0508 | 0.0497 | 0.0505 | 0.0515 | 0.0000 | 0.0000 | 0.0000 |
| 1993 | 0.0000 | 0.0004 | 0.0032 | 0.0084 | 0.0170 | 0.0244 | 0.0311 | 0.0363 | 0.0452 | 0.0624 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1994 | 0.0000 | 0.0007 | 0.0053 | 0.0139 | 0.0280 | 0.0404 | 0.0516 | 0.0603 | 0.0753 | 0.1045 | 0.1023 | 0.1038 | 0.1061 | 0.1375 | 0.1375 | 0.0000 |
| 1995 | 0.0000 | 0.0014 | 0.0109 | 0.0290 | 0.0589 | 0.0855 | 0.1100 | 0.1291 | 0.1627 | 0.2301 | 0.2248 | 0.2285 | 0.2337 | 0.3094 | 0.0000 | 0.3094 |
| 1996 | 0.0000 | 0.0018 | 0.0138 | 0.0367 | 0.0747 | 0.1090 | 0.1407 | 0.1656 | 0.2097 | 0.2996 | 0.2925 | 0.2974 | 0.3045 | 0.0000 | 0.0000 | 0.0000 |
| 1997 | 0.0000 | 0.0030 | 0.0233 | 0.0626 | 0.1295 | 0.1913 | 0.2502 | 0.2975 | 0.3843 | 0.5739 | 0.5582 | 0.5690 | 0.5847 | 0.8331 | 0.8331 | 0.0000 |
| 1998 | 0.0000 | 0.0024 | 0.0185 | 0.0496 | 0.1018 | 0.1493 | 0.1940 | 0.2295 | 0.2933 | 0.4275 | 0.4167 | 0.4242 | 0.4350 | 0.5985 | 0.5985 | 0.0000 |
| 1999 | 0.0000 | 0.0018 | 0.0138 | 0.0367 | 0.0747 | 0.1089 | 0.1406 | 0.1655 | 0.2096 | 0.2995 | 0.2924 | 0.2973 | 0.3044 | 0.4081 | 0.0000 | 0.4081 |
| 2000 | 0.0000 | 0.0020 | 0.0151 | 0.0402 | 0.0821 | 0.1198 | 0.1549 | 0.1826 | 0.2318 | 0.3330 | 0.3249 | 0.3305 | 0.3385 | 0.4567 | 0.4567 | 0.4567 |
| 2001 | 0.0000 | 0.0011 | 0.0084 | 0.0222 | 0.0449 | 0.0651 | 0.0834 | 0.0977 | 0.1226 | 0.1719 | 0.1681 | 0.1707 | 0.1745 | 0.2288 | 0.0000 | 0.2288 |
| 2002 | 0.0000 | 0.0014 | 0.0110 | 0.0292 | 0.0594 | 0.0863 | 0.1110 | 0.1303 | 0.1642 | 0.2323 | 0.2270 | 0.2307 | 0.2360 | 0.3125 | 0.3125 | 0.3125 |
| 2003 | 0.0000 | 0.0011 | 0.0084 | 0.0223 | 0.0452 | 0.0655 | 0.0840 | 0.0983 | 0.1234 | 0.1731 | 0.1692 | 0.1719 | 0.1757 | 0.2303 | 0.2303 | 0.0000 |

Table 3.4.1.4 NSS herring. ISVPA summary results - Norwegian spring spawning herring.

| year | B(1+) | SSB | R(1) | F(5-14, w-d by N(a)) |
| :---: | :---: | :---: | :---: | :---: |
| 1950 | 17654939 | 13087729 | 30138411 | 0.062 |
| 1951 | 19330620 | 12791718 | 321818228 | 0.052 |
| 1952 | 20864054 | 11446095 | 59433332 | 0.061 |
| 1953 | 18874740 | 10643050 | 30542849 | 0.067 |
| 1954 | 19532880 | 9281788 | 31362621 | 0.127 |
| 1955 | 16695288 | 10189255 | 10298732 | 0.100 |
| 1956 | 14624094 | 11993781 | 6854115 | 0.161 |
| 1957 | 11624893 | 10465208 | 8726573 | 0.137 |
| 1958 | 10178054 | 9590630 | 7137281 | 0.130 |
| 1959 | 8179188 | 7907567 | 3230913 | 0.151 |
| 1960 | 7834329 | 6419458 | 156311627 | 0.212 |
| 1961 | 8030732 | 4825066 | 72099785 | 0.126 |
| 1962 | 7265547 | 4065023 | 26987917 | 0.122 |
| 1963 | 7199450 | 3200784 | 5382250 | 0.164 |
| 1964 | 7233232 | 3477027 | 65619925 | 0.162 |
| 1965 | 6313123 | 3549972 | 35893563 | 0.387 |
| 1966 | 4394338 | 2940251 | 1992966 | 0.781 |
| 1967 | 2878785 | 1378546 | 18411013 | 1.371 |
| 1968 | 882363 | 259517 | 1374507 | 1.329 |
| 1969 | 118483 | 88390 | 972036 | 0.724 |
| 1970 | 86249 | 38367 | 4348950 | 0.936 |
| 1971 | 135049 | 12229 | 329396 | 0.440 |
| 1972 | 93787 | 3825 | 106156 | 0.609 |
| 1973 | 121493 | 105214 | 134959 | 0.210 |
| 1974 | 181322 | 123921 | 5181226 | 0.046 |
| 1975 | 328544 | 113353 | 3484678 | 0.031 |
| 1976 | 380820 | 171208 | 1181702 | 0.028 |
| 1977 | 445331 | 321860 | 3438676 | 0.035 |
| 1978 | 577341 | 393351 | 2137845 | 0.022 |
| 1979 | 632388 | 420964 | 2549826 | 0.018 |
| 1980 | 744878 | 492880 | 4839758 | 0.013 |
| 1981 | 784267 | 513912 | 617367 | 0.022 |
| 1982 | 717629 | 509712 | 465161 | 0.015 |
| 1983 | 751829 | 578310 | 907877 | 0.031 |
| 1984 | 1946065 | 602738 | 127343366 | 0.037 |
| 1985 | 4987381 | 509171 | 4898450 | 0.190 |
| 1986 | 1689898 | 418786 | 6735948 | 1.140 |
| 1987 | 2867516 | 859788 | 3815091 | 0.660 |
| 1988 | 3268714 | 2689645 | 3616859 | 0.179 |
| 1989 | 3752458 | 3238683 | 10739134 | 0.063 |
| 1990 | 4147921 | 3272179 | 28037640 | 0.065 |
| 1991 | 4634905 | 3398520 | 50065878 | 0.038 |
| 1992 | 5675250 | 3275232 | 137219228 | 0.034 |
| 1993 | 7348727 | 3201482 | 182796101 | 0.051 |
| 1994 | 8648050 | 3696603 | 49090778 | 0.068 |
| 1995 | 9697402 | 4717313 | 18243865 | 0.118 |
| 1996 | 9967246 | 6615146 | 5405047 | 0.110 |
| 1997 | 10390956 | 8412726 | 38016920 | 0.187 |
| 1998 | 9329120 | 7924784 | 33342379 | 0.163 |
| 1999 | 11535696 | 7656487 | 154995219 | 0.144 |
| 2000 | 11114284 | 6377770 | 113165751 | 0.191 |
| 2001 | 10718409 | 5923945 | 58193004 | 0.109 |
| 2002 | 11149780 | 6569109 | 15771098 | 0.156 |
| 2003 | 13418021 | 8681218 | 128782619 | 0.081 |


|  | 2003 | 2004 |
| :--- | :--- | :--- |
| Year classes in main tuning (main <br> surveys) | $1983,1990-1993,1996$ | $1983,1990-1993,1996-1999$ |
| Distributional assumption main surveys <br> (surveys 1,2,3 and 5) | Gamma, <br> CV | Gamma, <br> CV |
| Distributional assumption surveys in the <br> Barents Sea (survey 4 and 6) | Lognormal, <br> Standard deviation | Lognormal, <br> Standard deviation |
| Inclusion of larval data | Yes | Yes |
| Distributional assumption larval data | Gamma, <br> CV | Gamma, <br> CV |
| Inclusion of tagging data | Yes | Yes |
| Distributional assumption tagging data | Poisson | Poisson |
| Inclusion of zerogroup data | Yes | Yes |
| Distributional assumption zerogroup <br> data | Lognormal, <br> Standard deviation | Lognormal, <br> Standard deviation |

Table 3.6.4 Settings used in SeaStar for assessments in 2003 and 2004 - Norwegian spring spawning herring.

Table 3.6.1 Modelled historic abundance SeaStar. - Norwegian spring spawning herring.

|  |  |  |  | 3 |  | 5 |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 750.68 | 26.46 | 14.2 | 10.87 | 4.02 | 4.98 | 8.61 | 8.00 | 1.96 | 2.80 | 3.20 | 2.58 | 5.63 | 6.15 | 0.95 | 2. | 5.69 |
|  | 14 | 30 | 9.4 | 5.42 | 9.1 | 29 |  | . 90 | 6 |  | 2.3 |  |  | 4.6 |  |  |  |
|  | 96.64 | 58.46 | 117.9 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | 30.54 |  | . |  |  |  |  |  |  | 3.84 |  |  |  |  |  |  |
|  | 2.09 | 1. | 9.19 | 6.92 | 39.9 | 2.59 | 3. |  |  | 2.2 |  | 2. | 0. |  |  |  |  |
| 1955 | 24.97 | 10.30 | 8. 25 | 3.19 | 5.71 | 33. | 2. | 2.63 | 3.6 |  | 1.75 | 2.56 | 2.10 | 0.66 | . 01 |  |  |
|  | 29.86 | 6.85 |  | . 03 | 2.6 | 4.66 | 26 |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  | 20.45 |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  | 15.76 |  |  |  |  |  |  |  |  |
|  | 412 | 3.23 |  | 0.16 | 0.3 | 0.38 | 1.45 |  | 2.39 | 2. | 0. |  |  |  | 0.54 |  |  |
| 1960 | 197.51 | 156.2 | 1.18 | 0.25 | 0. | 0.31 | 0.3 | 1.1 | 1.0 | 1.8 | 9.8 | 0. | 0. | 0.94 | 0.33 |  |  |
| 1961 | 76 | 72.09 | 54.88 | 0.23 | 0.10 | 0.09 | 0.24 | 0.2 | 0.87 | 0.82 | 1.39 | 7.42 | 0.51 | 0.50 | 0.67 |  |  |
|  | 10. | 26.98 | 19.06 | 20.47 | 0.17 | 0.08 | 0.08 | 0.19 |  |  | 0.66 |  |  |  |  |  |  |
|  | 168 | 5.3 |  |  |  | 0.14 | 0.0 |  | 0.15 |  |  |  |  |  | 0.29 |  |  |
|  | 93.90 |  |  |  |  | 12.9 |  |  |  |  |  |  |  |  | 2.72 |  |  |
|  | 8. | 35 |  |  |  |  |  |  | 0. | 0. | 0. |  |  | 0.23 | 0.32 |  |  |
| 1966 | 51 | 1.98 | 12 | 8.32 | 0.0 | 1.23 | 3.1 | 5.90 | 0.06 | 0.02 | 0.03 |  | 0.03 | 0.10 | 0.10 |  |  |
|  | 3.95 | 18 | 0.38 | 3.88 | 5.26 | 0.05 | 0.6 | 1.50 | 2.40 |  |  |  |  |  |  |  |  |
|  | 5.19 | 1.33 |  |  | 2.05 | 1.5 | 0.02 | 0.1 | 0.24 |  | 0.00 |  |  |  |  |  |  |
| 1969 |  |  |  | 0.23 | 0.0 | 0.02 | 0.0 | 0.00 | 0.0 | 0.08 |  |  | 0.00 | 0.00 | 0.00 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0.24 |  |  |  |  | 0.01 | 0.0 |  | 0.00 |  |  |  |  |  | 0.00 |  |  |
|  | 0 |  |  |  | 0.0 |  | 0.00 |  |  | 0.00 |  |  |  |  |  |  |  |
|  | 12 | 0.17 |  |  | 0.3 | 0.00 | 0.00 | 0.00 | 0.00 |  |  |  |  |  |  |  |  |
|  | 8.63 |  |  |  | 0.0 |  | 0.0 |  | 0.00 |  |  |  |  |  |  |  |  |
|  | 2.9 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 10 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 5.10 |  |  |  |  | 0.01 | 0. |  |  | 0.00 |  |  |  |  | 0.00 |  | 0.73 |
|  | 6.20 | 2.0 | 1.6 | 0.19 | 0.4 | 0.60 | 0.0 | 0.00 | 0.00 | 0.08 |  |  |  |  | 0.00 |  |  |
|  | 12 | 2.5 |  |  | 0.1 | 0.3 | 0.4 |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  | 0.3 |  |  |  |  |  |  |  |  |  |  |
|  | 1.1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 2. |  |  | 0.83 | 0.3 | 0.24 |  | 0.10 | 0.2 |  |  |  |  |  | 0.00 |  |  |
| 3 | 334.36 | 0.9 | 0. | 0.10 | 0. | 0.3 | 0. | 0.35 | 0.0 | 0.20 |  |  | 0. | 0.00 | 0.03 | 0. | 0.3 |
| 198 | 11.53 | 135.86 | 0.3 | 0.0 | 0.0 | 0.58 | 0. | 0.17 | 0.2 | 0.0 | 0. |  | 0. | 0.00 | 0.00 | 0.0 |  |
| 19 | 35.7 | 4.6 | 55. | 0.15 | 0.0 | 0.06 | 0. | 0.19 | 0. | 0.24 | 0.05 |  |  | 0.0 | 0.00 |  |  |
| 19 | 6.0 | 14.5 | 1.89 | 22.33 | 0. | 0. | 0.0 | 0.26 | 0. | 0.06 | 0. |  |  | 0.10 | 0.00 |  |  |
| 19 | 8.99 | 2.45 | 5.9 | 0.7 | 18.7 | 0.08 | 0.0 | 0. | 0. | 0. | 0. | 0. | 0. | 0.01 | 0.01 |  |  |
| 1988 | 25.3 | 3.6 | 0.9 | 2.37 | 0.6 | 15.64 | 0.0 | 0. | 0.0 | 0.08 |  |  | 0.04 | 0.00 | 0.00 | . 00 |  |
| 989 | 70.0 | 10.3 | 1.48 | 0.40 | 1.99 | 0.53 | 12.95 | 0.0 | 0.0 | 0.00 | 0.06 | 0.00 | 0.0 | 0.03 | 0.00 |  |  |

Table 3.6.1. continued on next page

Table 3.6.1 continued.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1990 | 122.95 | 28.46 | 4.19 | 0.59 | 0.34 | 1.71 | 0.45 | 10.85 | 0.03 | 0.00 | 0.00 | 0.05 | 0.00 | 0.00 | 0.03 | 0.00 | 0.11 |
| 1991 | 331.07 | 49.99 | 11.57 | 1.69 | 0.49 | 0.29 | 1.46 | 0.38 | 9.13 | 0.02 | 0.00 | 0.00 | 0.04 | 0.00 | 0.00 | 0.02 | 0.10 |
| 1992 | 397.93 | 134.60 | 20.32 | 4.70 | 1.45 | 0.42 | 0.25 | 1.24 | 0.32 | 7.65 | 0.02 | 0.00 | 0.00 | 0.03 | 0.00 | 0.00 | 0.11 |
| 1993 | 96.53 | 161.78 | 54.72 | 8.26 | 4.04 | 1.22 | 0.35 | 0.21 | 1.06 | 0.27 | 6.38 | 0.01 | 0.00 | 0.00 | 0.03 | 0.00 | 0.09 |
| 1994 | 32.48 | 39.24 | 65.78 | 22.24 | 7.08 | 3.37 | 0.97 | 0.30 | 0.18 | 0.88 | 0.21 | 5.11 | 0.01 | 0.00 | 0.00 | 0.02 | 0.08 |
| 1995 | 9.08 | 13.20 | 15.95 | 26.74 | 19.12 | 5.99 | 2.57 | 0.68 | 0.24 | 0.15 | 0.72 | 0.15 | 3.80 | 0.01 | 0.00 | 0.00 | 0.09 |
| 1996 | 66.32 | 3.69 | 5.37 | 6.49 | 22.96 | 16.13 | 4.58 | 1.62 | 0.37 | 0.19 | 0.11 | 0.56 | 0.05 | 2.42 | 0.00 | 0.00 | 0.08 |
| 1997 | 38.96 | 26.96 | 1.50 | 2.16 | 5.55 | 19.10 | 12.43 | 3.07 | 1.02 | 0.22 | 0.16 | 0.09 | 0.42 | 0.03 | 1.31 | 0.00 | 0.06 |
| 1998 | 264.99 | 15.84 | 10.96 | 0.60 | 1.74 | 4.53 | 14.77 | 8.85 | 1.94 | 0.57 | 0.14 | 0.12 | 0.05 | 0.28 | 0.01 | 0.78 | 0.06 |
| 1999 | 202.12 | 107.74 | 6.44 | 4.40 | 0.45 | 1.27 | 3.55 | 11.08 | 6.44 | 1.31 | 0.37 | 0.08 | 0.08 | 0.04 | 0.13 | 0.00 | 0.66 |
| 2000 | 12.71 | 82.18 | 43.80 | 2.62 | 3.66 | 0.35 | 0.97 | 2.66 | 8.05 | 4.46 | 0.86 | 0.22 | 0.05 | 0.03 | 0.03 | 0.03 | 0.51 |
| 2001 | 24.16 | 5.17 | 33.41 | 17.80 | 2.17 | 2.63 | 0.27 | 0.73 | 1.91 | 5.72 | 2.87 | 0.54 | 0.12 | 0.03 | 0.01 | 0.00 | 0.45 |
| 2002 | 202.84 | 9.82 | 2.10 | 13.58 | 15.23 | 1.72 | 1.87 | 0.20 | 0.54 | 1.37 | 4.15 | 2.00 | 0.40 | 0.08 | 0.02 | 0.00 | 0.38 |
| 2003 | 170.68 | 82.47 | 3.99 | 0.81 | 11.51 | 12.51 | 1.24 | 1.31 | 0.14 | 0.38 | 0.94 | 2.95 | 1.41 | 0.29 | 0.06 | 0.01 | 0.32 |

Table 3.6.2 Modelled historic F-values SeaStar - Norwegian spring spawning herring..
$\begin{array}{llllllllllllllllll}0 & 1 & 2 & 3 & 4 & 5 & 6 & 7 & 8 & 9 & 10 & 11 & 12 & 13 & 14 & 15 & 16\end{array}$ $1 \begin{array}{lllllllllllllllll}1950 & 0.011 & 0.126 & 0.068 & 0.028 & 0.051 & 0.041 & 0.071 & 0.088 & 0.045 & 0.035 & 0.038 & 0.037 & 0.038 & 0.067 & 0.078 & 0.046\end{array} 0.046$ $\begin{array}{lllllllllllllllllll}1951 & 0.018 & 0.040 & 0.068 & 0.001 & 0.047 & 0.058 & 0.044 & 0.084 & 0.109 & 0.053 & 0.039 & 0.043 & 0.056 & 0.060 & 0.079 & 0.070 & 0.070\end{array}$ $\begin{array}{lllllllllllllllll}1952 & 0.252 & 0.282 & 0.017 & 0.012 & 0.014 & 0.091 & 0.057 & 0.067 & 0.078 & 0.087 & 0.067 & 0.049 & 0.054 & 0.068 & 0.055 & 0.073\end{array} 0.073$ $\begin{array}{lllllllllllllllllll}1953 & 0.110 & 0.301 & 0.052 & 0.017 & 0.017 & 0.028 & 0.067 & 0.041 & 0.045 & 0.081 & 0.117 & 0.065 & 0.064 & 0.058 & 0.079 & 0.078 & 0.078\end{array}$ $\begin{array}{llllllllllllllllll}1954 & 0.507 & 0.436 & 0.158 & 0.042 & 0.040 & 0.061 & 0.080 & 0.118 & 0.080 & 0.101 & 0.148 & 0.185 & 0.118 & 0.089 & 0.104 & 0.130 & 0.130\end{array}$ $\begin{array}{llllllllllllllllll}1955 & 0.393 & 0.575 & 0.102 & 0.032 & 0.054 & 0.069 & 0.060 & 0.081 & 0.085 & 0.067 & 0.127 & 0.133 & 0.110 & 0.101 & 0.095 & 0.104 & 0.104\end{array}$ $\begin{array}{llllllllllllllllll}1956 & 0.331 & 0.622 & 0.540 & 0.042 & 0.107 & 0.075 & 0.110 & 0.072 & 0.111 & 0.105 & 0.131 & 0.176 & 0.165 & 0.115 & 0.143 & 0.130 & 0.130\end{array}$ $\begin{array}{lllllllllllllllll}1957 & 0.369 & 0.896 & 0.262 & 0.046 & 0.176 & 0.084 & 0.069 & 0.111 & 0.059 & 0.089 & 0.093 & 0.116 & 0.147 & 0.121 & 0.121 & 0.100\end{array} 0.100$ $\begin{array}{llllllllllllllllll}1958 & 1.069 & 0.954 & 1.282 & 0.041 & 0.043 & 0.069 & 0.061 & 0.073 & 0.069 & 0.071 & 0.111 & 0.133 & 0.174 & 0.125 & 0.074 & 0.082 & 0.082\end{array}$ $\begin{array}{lllllllllllllllll}1959 & 0.070 & 0.102 & 0.611 & 0.105 & 0.078 & 0.077 & 0.115 & 0.099 & 0.115 & 0.099 & 0.114 & 0.148 & 0.178 & 0.221 & 0.168 & 0.113\end{array} 0.113$ $\begin{array}{lllllllllllllllll}1960 & 0.108 & 0.147 & 0.733 & 0.759 & 0.168 & 0.104 & 0.091 & 0.098 & 0.080 & 0.128 & 0.136 & 0.145 & 0.214 & 0.193 & 0.204 & 0.139\end{array} 0.139$ $\begin{array}{llllllllllllllllll}1961 & 0.137 & 0.430 & 0.086 & 0.157 & 0.092 & 0.049 & 0.070 & 0.092 & 0.080 & 0.067 & 0.112 & 0.112 & 0.112 & 0.101 & 0.107 & 0.106 & 0.106\end{array}$ $\begin{array}{llllllllllllllllll}1962 & 0.364 & 0.271 & 0.090 & 0.102 & 0.052 & 0.044 & 0.109 & 0.121 & 0.071 & 0.097 & 0.090 & 0.125 & 0.167 & 0.130 & 0.163 & 0.147 & 0.147\end{array}$ $\begin{array}{llllllllllllllllll}1963 & 0.046 & 0.964 & 0.483 & 0.123 & 0.058 & 0.042 & 0.031 & 0.069 & 0.145 & 0.069 & 0.243 & 0.214 & 0.263 & 0.274 & 0.344 & 0.257 & 0.257\end{array}$ $\begin{array}{llllllllllllllllll}1964 & 0.062 & 0.067 & 0.536 & 0.061 & 0.083 & 0.187 & 0.138 & 0.031 & 0.072 & 0.283 & 0.306 & 0.333 & 0.284 & 0.366 & 0.365 & 0.320 & 0.320\end{array}$ $1 \begin{array}{llllllllllllllll}1965 & 0.554 & 0.181 & 0.198 & 0.672 & 0.177 & 0.156 & 0.297 & 0.280 & 0.444 & 0.239 & 0.344 & 0.838 & 0.659 & 0.691 & 0.627 \\ 0.620 & 0.620\end{array}$ $\begin{array}{llllllllllllllllll}1966 & 0.128 & 0.742 & 0.244 & 0.308 & 0.405 & 0.529 & 0.593 & 0.749 & 1.307 & 0.976 & 1.343 & 1.064 & 0.543 & 1.370 & 1.484 & 1.188 & 1.188\end{array}$ $\begin{array}{llllllllllllllllll}1967 & 0.186 & 1.845 & 0.339 & 0.489 & 1.099 & 0.853 & 1.311 & 1.690 & 1.483 & 1.328 & 1.451 & 1.082 & 1.257 & 1.131 & 1.992 & 1.478 & 1.478\end{array}$ $\begin{array}{lllllllllllllllllll}1968 & 0.775 & 0.722 & 0.724 & 3.232 & 4.591 & 4.749 & 1.812 & 1.214 & 0.936 & 1.577 & 1.288 & 1.320 & 0.771 & 1.964 & 1.245 & 1.362 & 1.362\end{array}$ $\begin{array}{llllllllllllllllll}1969 & 0.094 & 1.707 & 1.867 & 2.046 & 0.259 & 0.756 & 0.599 & 0.347 & 0.419 & 0.600 & 0.626 & 0.612 & 0.867 & 0.394 & 0.843 & 0.578 & 0.578\end{array}$ $\begin{array}{llllllllllllllllll}1970 & 0.333 & 0.261 & 1.297 & 0.529 & 1.486 & 0.298 & 0.678 & 1.105 & 1.164 & 1.178 & 1.368 & 1.555 & 3.941 & 1.618 & 0.823 & 1.406 & 1.406\end{array}$ $\begin{array}{llllllllllllllllll}1971 & 0.226 & 0.429 & 0.125 & 0.283 & 0.140 & 0.308 & 0.277 & 0.476 & 1.626 & 2.550 & 1.958 & 2.279 & 1.841 & 0.217 & 2.804 & 2.028 & 2.028\end{array}$ $\begin{array}{llllllllllllllllll}1972 & 0.842 & 1.834 & 0.987 & 0.098 & 1.293 & 0.956 & 1.819 & 1.252 & 2.980 & 1.779 & 0.039 & 1.571 & 2.069 & 1.110 & 0.332 & 2.106 & 2.106\end{array}$ $\begin{array}{llllllllllllllllll}1973 & 0.004 & 0.033 & 0.769 & 0.405 & 0.089 & 0.857 & 1.521 & 1.117 & 1.365 & 0.015 & 0.029 & 0.047 & 0.009 & 0.014 & 1.264 & 0.603 & 0.603\end{array}$ $\begin{array}{llllllllllllllllll}1974 & 0.012 & 0.002 & 0.097 & 0.122 & 0.060 & 0.112 & 0.973 & 0.779 & 0.009 & 0.019 & 0.017 & 0.035 & 0.058 & 0.011 & 0.016 & 0.017 & 0.017\end{array}$ $\begin{array}{lllllllllllllllllll}1975 & 0.016 & 0.002 & 0.001 & 0.157 & 0.222 & 0.319 & 0.189 & 0.038 & 0.014 & 0.010 & 0.023 & 0.021 & 0.042 & 0.072 & 0.012 & 0.019 & 0.019\end{array}$ $\begin{array}{llllllllllllllllllll}1976 & 0.003 & 0.003 & 0.001 & 0.030 & 0.397 & 0.002 & 0.000 & 0.109 & 0.009 & 0.008 & 0.012 & 0.027 & 0.024 & 0.052 & 0.090 & 0.015 & 0.015\end{array}$ $\begin{array}{lllllllllllllllllll}1977 & 0.013 & 0.002 & 0.010 & 0.043 & 0.036 & 0.036 & 0.003 & 0.266 & 0.116 & 0.011 & 0.010 & 0.014 & 0.033 & 0.029 & 0.064 & 0.116 & 0.116\end{array}$ $\begin{array}{lllllllllllllllllll}1978 & 0.005 & 0.002 & 0.001 & 0.017 & 0.028 & 0.037 & 0.115 & 0.003 & 0.743 & 0.069 & 0.012 & 0.011 & 0.017 & 0.039 & 0.035 & 0.079 & 0.079\end{array}$ $\begin{array}{llllllllllllllllllll}1979 & 0.004 & 0.002 & 0.004 & 0.010 & 0.012 & 0.019 & 0.025 & 0.055 & 0.003 & 0.002 & 0.043 & 0.015 & 0.013 & 0.020 & 0.048 & 0.042 & 0.042\end{array}$ $\begin{array}{llllllllllllllllllll}1980 & 0.007 & 0.000 & 0.001 & 0.021 & 0.011 & 0.018 & 0.027 & 0.042 & 0.093 & 0.033 & 0.002 & 0.055 & 0.017 & 0.016 & 0.023 & 0.058 & 0.058\end{array}$ $\begin{array}{lllllllllllllllllll}1981 & 0.012 & 0.003 & 0.009 & 0.011 & 0.018 & 0.019 & 0.020 & 0.018 & 0.026 & 0.092 & 0.692 & 0.384 & 0.024 & 0.020 & 0.019 & 0.028 & 0.028\end{array}$ $\begin{array}{llllllllllllllllllll}1982 & 0.015 & 0.004 & 0.001 & 0.018 & 0.024 & 0.020 & 0.017 & 0.021 & 0.023 & 0.023 & 0.040 & 0.539 & 0.194 & 0.004 & 0.024 & 0.022 & 0.022\end{array}$ $\begin{array}{llllllllllllllllllll}1983 & 0.001 & 0.008 & 0.015 & 0.036 & 0.033 & 0.035 & 0.034 & 0.021 & 0.017 & 0.025 & 0.033 & 0.058 & 2.280 & 3.311 & 0.030 & 0.029 & 0.029\end{array}$ $\begin{array}{lllllllllllllllllll}1984 & 0.005 & 0.000 & 0.010 & 0.070 & 0.074 & 0.121 & 0.083 & 0.084 & 0.059 & 0.116 & 0.112 & 0.035 & 0.000 & 0.294 & 0.242 & 0.070 & 0.070\end{array}$ $\begin{array}{llllllllllllllllllll}1985 & 0.001 & 0.004 & 0.006 & 0.165 & 0.343 & 0.320 & 0.378 & 0.396 & 0.590 & 0.336 & 0.221 & 0.302 & 0.389 & 0.001 & 0.506 & 0.379 & 0.379\end{array}$ $\begin{array}{llllllllllllllllll}1986 & 0.004 & 0.000 & 0.003 & 0.026 & 0.187 & 0.588 & 0.530 & 0.565 & 1.264 & 1.250 & 0.835 & 0.820 & 2.057 & 2.161 & 0.001 & 1.396 & 1.396\end{array}$ $\begin{array}{llllllllllllllllll}1987 & 0.002 & 0.004 & 0.010 & 0.028 & 0.029 & 0.293 & 0.254 & 0.461 & 0.268 & 0.639 & 1.062 & 0.093 & 0.903 & 1.516 & 1.516 & 0.417 & 0.417\end{array}$ $\begin{array}{llllllllllllllllll}1988 & 0.001 & 0.001 & 0.015 & 0.029 & 0.043 & 0.039 & 0.223 & 0.436 & 0.857 & 0.206 & 1.464 & 1.056 & 0.088 & 0.875 & 4.757 & 0.381 & 0.381\end{array}$ $\begin{array}{llllllllllllllllllllll}1989 & 0.000 & 0.000 & 0.027 & 0.008 & 0.002 & 0.012 & 0.027 & 0.113 & 0.149 & 0.198 & 0.062 & 0.912 & 0.744 & 0.010 & 0.171 & 0.090 & 0.090\end{array}$

Table 3.6.2 Continued on next page

Table 3.6.2 Continued

| 1990 | 0.000 | 0.000 | 0.006 | 0.035 | 0.008 | 0.008 | 0.026 | 0.023 | 0.053 | 0.436 | 1.430 | 0.056 | 1.678 | 0.413 | 0.022 | 0.153 | 0.153 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1991 | 0.000 | 0.000 | 0.000 | 0.005 | 0.006 | 0.005 | 0.011 | 0.026 | 0.026 | 0.130 | 0.214 | 0.172 | 0.019 | 1.646 | 1.990 | 0.024 | 0.024 |
| 1992 | 0.000 | 0.000 | 0.000 | 0.003 | 0.025 | 0.013 | 0.005 | 0.010 | 0.020 | 0.032 | 0.175 | 0.486 | 0.963 | 0.041 | 0.048 | 0.029 | 0.029 |
| 1993 | 0.000 | 0.000 | 0.000 | 0.004 | 0.029 | 0.080 | 0.027 | 0.019 | 0.031 | 0.078 | 0.072 | 0.000 | 0.001 | 0.008 | 0.000 | 0.059 | 0.059 |
| 1994 | 0.000 | 0.000 | 0.000 | 0.002 | 0.017 | 0.124 | 0.203 | 0.058 | 0.050 | 0.047 | 0.200 | 0.146 | 0.349 | 0.939 | 2.220 | 0.100 | 0.100 |
| 1995 | 0.000 | 0.000 | 0.000 | 0.002 | 0.020 | 0.119 | 0.312 | 0.456 | 0.072 | 0.124 | 0.110 | 0.925 | 0.299 | 1.200 | 4.053 | 0.100 | 0.100 |
| 1996 | 0.000 | 0.000 | 0.009 | 0.006 | 0.034 | 0.111 | 0.250 | 0.316 | 0.358 | 0.032 | 0.074 | 0.136 | 0.463 | 0.465 | 0.001 | 0.303 | 0.303 |
| 1997 | 0.000 | 0.000 | 0.023 | 0.067 | 0.054 | 0.107 | 0.190 | 0.311 | 0.426 | 0.348 | 0.144 | 0.497 | 0.265 | 1.365 | 0.364 | 0.262 | 0.262 |
| 1998 | 0.000 | 0.000 | 0.012 | 0.136 | 0.163 | 0.092 | 0.137 | 0.167 | 0.239 | 0.282 | 0.413 | 0.259 | 0.084 | 0.576 | 6.599 | 0.161 | 0.161 |
| 1999 | 0.000 | 0.000 | 0.001 | 0.034 | 0.090 | 0.121 | 0.140 | 0.170 | 0.217 | 0.273 | 0.369 | 0.227 | 0.780 | 0.236 | 1.245 | 0.163 | 0.163 |
| 2000 | 0.000 | 0.000 | 0.001 | 0.035 | 0.180 | 0.113 | 0.131 | 0.179 | 0.191 | 0.291 | 0.317 | 0.430 | 0.402 | 1.505 | 8.044 | 0.166 | 0.166 |
| 2001 | 0.000 | 0.000 | 0.000 | 0.006 | 0.083 | 0.192 | 0.168 | 0.152 | 0.183 | 0.172 | 0.211 | 0.159 | 0.232 | 0.132 | 0.921 | 0.168 | 0.168 |
| 2002 | 0.000 | 0.000 | 0.047 | 0.016 | 0.047 | 0.174 | 0.208 | 0.178 | 0.206 | 0.233 | 0.189 | 0.202 | 0.156 | 0.173 | 0.398 | 0.170 | 0.170 |
| 2003 | 0.000 | 0.000 | 0.002 | 0.104 | 0.031 | 0.067 | 0.171 | 0.153 | 0.200 | 0.247 | 0.294 | 0.236 | 0.187 | 0.177 | 0.175 | 0.173 | 0.173 |

Table 3.6.3 Summary table SeaStar. Biomass in million tonnes - Norwegian spring spawning herring..

|  | $\begin{aligned} & \text { Recruits } \\ & \text { Age } 0 \end{aligned}$ | Total <br> biomass | Spawning stock biomass | $\begin{gathered} \text { Fbar } \\ 5-14 \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: |
| 1950 | 750.680 | 19.642 | 13.992 | 0.058 |
| 1951 | 146.355 | 19.798 | 13.145 | 0.070 |
| 1952 | 96.644 | 21.187 | 12.028 | 0.073 |
| 1953 | 86.102 | 18.945 | 10.949 | 0.066 |
| 1954 | 42.086 | 20.508 | 10.593 | 0.113 |
| 1955 | 24.971 | 17.830 | 11.383 | 0.078 |
| 1956 | 29.858 | 15.933 | 13.016 | 0.110 |
| 1957 | 25.397 | 13.161 | 11.686 | 0.103 |
| 1958 | 23.094 | 11.493 | 10.634 | 0.079 |
| 1959 | 412.478 | 9.893 | 8.969 | 0.113 |
| 1960 | 197.514 | 9.460 | 7.626 | 0.136 |
| 1961 | 76.103 | 9.461 | 6.024 | 0.104 |
| 1962 | 19.003 | 8.246 | 4.880 | 0.146 |
| 1963 | 168.931 | 8.346 | 4.015 | 0.253 |
| 1964 | 93.903 | 7.838 | 3.814 | 0.226 |
| 1965 | 8.491 | 7.147 | 4.120 | 0.278 |
| 1966 | 51.409 | 5.552 | 3.650 | 0.696 |
| 1967 | 3.947 | 4.183 | 2.273 | 1.519 |
| 1968 | 5.187 | 1.943 | 1.131 | 3.493 |
| 1969 | 9.785 | 1.087 | 0.972 | 0.590 |
| 1970 | 0.661 | 0.892 | 0.729 | 1.320 |
| 1971 | 0.236 | 0.807 | 0.572 | 1.525 |
| 1972 | 0.957 | 0.673 | 0.483 | 1.497 |
| 1973 | 12.884 | 0.766 | 0.692 | 1.173 |
| 1974 | 8.631 | 0.706 | 0.628 | 0.114 |
| 1975 | 2.971 | 0.773 | 0.542 | 0.190 |
| 1976 | 10.068 | 0.768 | 0.539 | 0.106 |
| 1977 | 5.095 | 0.780 | 0.630 | 0.111 |
| 1978 | 6.201 | 0.892 | 0.667 | 0.043 |
| 1979 | 12.498 | 0.905 | 0.656 | 0.024 |
| 1980 | 1.474 | 0.981 | 0.702 | 0.034 |
| 1981 | 1.100 | 1.002 | 0.710 | 0.022 |
| 1982 | 2.343 | 0.904 | 0.679 | 0.020 |
| 1983 | 334.362 | 1.245 | 0.723 | 0.029 |
| 1984 | 11.528 | 2.170 | 0.712 | 0.090 |
| 1985 | 35.722 | 5.432 | 0.593 | 0.379 |
| 1986 | 6.041 | 1.939 | 0.472 | 1.074 |
| 1987 | 8.995 | 3.287 | 0.968 | 0.404 |
| 1988 | 25.393 | 3.664 | 2.896 | 0.043 |
| 1989 | 70.008 | 4.260 | 3.515 | 0.027 |
| 1990 | 122.947 | 4.766 | 3.667 | 0.022 |
| 1991 | 331.066 | 5.435 | 3.804 | 0.024 |
| 1992 | 397.925 | 6.509 | 3.667 | 0.028 |
| 1993 | 96.526 | 7.637 | 3.508 | 0.065 |
| 1994 | 32.477 | 8.714 | 3.963 | 0.134 |
| 1995 | 9.076 | 9.572 | 4.859 | 0.225 |
| 1996 | 66.315 | 9.685 | 6.496 | 0.185 |
| 1997 | 38.959 | 9.677 | 7.842 | 0.173 |
| 1998 | 264.987 | 8.475 | 7.122 | 0.155 |
| 1999 | 202.118 | 9.642 | 6.583 | 0.192 |
| 2000 | 12.705 | 8.746 | 5.293 | 0.233 |
| 2001 | 24.157 | 7.352 | 4.577 | 0.184 |
| 2002 | 202.841 | 7.795 | 4.594 | 0.197 |
| 2003 | 170.677 | 8.794 | 5.787 | 0.126 |

Table 3.6.1.1. Norwegian spring-spawning herring, input data to the short term prediction, SeaStar.
Landings in 2004, 0.825 million tonnes
Fbar age range: $5-14$, Fbar is weighted with population numbers January 1
Spawning stock biomass in 2004, 6.30015 million tonnes
Fbar in 2004, 0.16583
Weight at age in stock and catch are assumed equal
Basis for weight in stock prediction: predicted condition factor
Basis for weight in catch prediction: mean of 2000-2002
Part of year before spawning: 0.1

| Age | Numbers <br> (billion) | Weight <br> stock (kg) | Weight <br> catch (kg) | Weight <br> catch $(\mathrm{kg})$ | Fraction <br> mature | Expl. <br> pattern |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 2004 | 2004 | 2005 |  |  |  |
| 0 | 0 | 0.001 | 0.001 | 0.007 | 0 | 0 |
| 1 | 69.392 | 0.01 | 0.01 | 0.063 | 0 | 0 |
| 2 | 33.527 | 0.037 | 0.037 | 0.094 | 0 | 0.002 |
| 3 | 1.621 | 0.055 | 0.055 | 0.171 | 0 | 0.104 |
| 4 | 0.632 | 0.212 | 0.214 | 0.218 | 0.45 | 0.031 |
| 5 | 9.601 | 0.241 | 0.243 | 0.257 | 0.9 | 0.067 |
| 6 | 10.069 | 0.279 | 0.281 | 0.288 | 1 | 0.171 |
| 7 | 0.902 | 0.302 | 0.304 | 0.319 | 1 | 0.153 |
| 8 | 0.964 | 0.337 | 0.34 | 0.327 | 1 | 0.2 |
| 9 | 0.1 | 0.354 | 0.357 | 0.354 | 1 | 0.247 |
| 10 | 0.255 | 0.355 | 0.358 | 0.363 | 1 | 0.294 |
| 11 | 0.6 | 0.36 | 0.363 | 0.361 | 1 | 0.236 |
| 12 | 2.006 | 0.371 | 0.374 | 0.381 | 1 | 0.187 |
| 13 | 1.006 | 0.4 | 0.403 | 0.397 | 1 | 0.177 |
| 14 | 0.21 | 0.412 | 0.415 | 0.402 | 1 | 0.175 |
| 15 | 0.044 | 0.445 | 0.448 | 0.45 | 1 | 0.173 |
| 16 | 0.284 | 0.445 | 0.448 | 0.45 | 1 | 0.173 |

Table 3.6.1.2. Norwegian spring-spawning herring, short term prediction. Total biomass is not included because the recruitment was set to 0 in the calculations.

Norwegian spring-spawning herring, WG 2004
Date and time: 04-05-2004 at 14:00
Fbar age range: $5-14$, Fbar is weighted with population numbers January 1
Basis for 2004: TAC constrained; recruitment: no recruitment in the prediction

| SSB | $\begin{gathered} 2004 \\ \text { FMult } \end{gathered}$ | Fbar | Landings | SSB | $\begin{gathered} 2005 \\ \text { FMult } \end{gathered}$ | Fbar | Landings | $\begin{gathered} 2006 \\ \mathrm{SSB} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6.30015 | 1 | 0.16583 | 0.825 | 6.391 | 0.51 | 0.084 | 0.609 | 6.667 |
|  |  |  |  | 6.380 | 0.6 | 0.100 | 0.725 | 6.556 |
|  |  |  |  | 6.370 | 0.71 | 0.117 | 0.839 | 6.446 |
|  |  |  |  | 6.365 | 0.75 | 0.125 | 0.892 | 6.396 |
|  |  |  |  | 6.359 | 0.81 | 0.134 | 0.952 | 6.339 |
|  |  |  |  | 6.348 | 0.9 | 0.150 | 1.063 | 6.233 |
|  |  |  |  | 6.338 | 1.01 | 0.167 | 1.172 | 6.130 |
|  |  |  |  | 6.327 | 1.11 | 0.184 | 1.279 | 6.028 |
|  |  |  |  | 6.317 | 1.21 | 0.201 | 1.385 | 5.928 |
|  |  |  |  | 6.306 | 1.31 | 0.217 | 1.490 | 5.830 |
|  |  |  |  | 6.296 | 1.41 | 0.234 | 1.592 | 5.734 |
|  |  |  |  | 6.285 | 1.51 | 0.251 | 1.694 | 5.639 |

Input units are millions and kg ; output in kilotonnes


Figure 3.1.3.1. Total catches of Norwegian spring-spawning herring in 2003 by ICES rectangle. Grading of the symbols: black dots less than 300 t , open squares 300-3 000 t , and black squares $>3000 \mathrm{t}$.


## Quarter 1



## Quarter 3



## Quarter 2



## Quarter 4

Figure 3.1.3.2. Total catches of Norwegian spring-spawning herring in 2003 by quarter and ICES rectangle. Grading of the symbols: black dots less than 300 t , open squares $300-3000 \mathrm{t}$, and black squares $>3000 \mathrm{t}$.


Figure 3.3.1. Comparison of perceived number by age in 2003 in SeaStar 2003 and 2004 assessments - Norwegian spring spawning herring.


Figure 3.3.2 Spawning stock timeseries for SeaStar runs - Norwegian spring spawning herring..


## NSS Herring.

ISVPA, catch-controlled version

## 0 - catch-at-age

1- spawning grounds acoustic in Febr.-March
2-acoust. surv. in wint. area Nov.-December
3 - acust. in wintering areas, January
4- Young herring in the Barents Sea (June)
5- Feeding areas, May
6-Young herring in the Barents Sea, September survey
Figure 3.4.1.1 Profiles of components of ISVPA loss function for different sources of data. Survey N5 is used in form of age proportions - Norwegian spring spawning herring..


Figure 3.4.1.2 Profiles of components of ISVPA loss function for different sources of data. Survey N5 is used as abundance-at-age index - Norwegian spring spawning herring.


Figure 3.4.1.3 Profiles of components of ISVPA loss function for different sources of data. Survey N5 is used in form of age proportions, years of run: 1986-2003 - Norwegian spring spawning herring..


Figure 3.4.1.4 NSS herring. ISVPA. Norwegian spring spawning herring. Comparison of runs:

1- all data, years: 1950-2003, survey N5-as age proportions
2 - no surveys (only catch-at-age is used), years: 1950-2003

3 - all data, years: 1986-2003, survey N5 - as age proportions

4 - all data, years: 1950-2003, survey N5 - as abundance index

5 - all data, years: 1950-2003, survey N 5 - as abundance index, first age of survey N 5 is 5 .

6 - all data, years: 1950-2003, survey N5 - as age proportions, first age of survey N5 is 5 .


Figure 3.4.1.5 NSS Herring. ISVPA. Plots of residuals - Norwegian spring spawning herring..


Figure 3.4.6.1 Estimated ISVPA selection patterns


Figure 3.4.1.7 NSS herring. ISVPA summary results - Norwegian spring spawning herring.




Figure 3.4.1.8 NSS herring. ISVPA. Retrospective runs - Norwegian spring spawning herring


Figure 3.5.2.1 Profiles of components of ISVPA loss function for different sources of data. Survey N5 is used as abundance-at-age index. Norwegian spring spawning herring.

Spawning stock
million tonnes

## ISVPA SeaStar comparisons



Figure 3.5.3 Spawning stock trajectories for Norwegian spring spawning herring for four different runs of SeaStar and ISVPA.




Figure 3.6.1 Fit of the VPA scaled with the estimated catchability to the observations in the survey on the spawning grounds for the tuned year classes - Norwegian spring spawning herring.




Figure 3.6.2 Fit of the VPA scaled with the estimated catchability to the observations in the December survey in Vestfjorden - Norwegian spring spawning herring.


Figure 3.6.3 Fit of the VPA scaled with the estimated catchability to the observations in the January survey in Vestfjorden - Norwegian spring spawning herring.



Figure 3.6.4 Fit of the VPA scaled with the estimated catchability to the observations in the May survey in the Barents Sea - Norwegian spring spawning herring.


Figure 3.6.5 Fit of the VPA scaled with the estimated catchability to the observations in the feeding areas in the Norwegian Sea - Norwegian spring spawning herring.


Figure 3.6.6. Fit of the VPA scaled with the estimated catchability to the observations for the Barents Sea September survey - Norwegian spring spawning herring..


Figure 3.6.7 Retrospective plot from SeaStar for Norwegian spring spawning herring.


Figure 3.6.8 Histogram of perceived SSB 2004 from SeaStar bootstrap runs - Norwegian spring spawning herring.

| F83 |  |
| :---: | :---: |
| F90 | $\because$ |
|  | $\text { 若 } \because$ |
| F91 |  |
| F92 |  |
| F93 | $\therefore$ |
| F96 |  |
| F97 |  |
| F98 |  |
| F99 |  |
| F00 | ( $\because \%$ |
| F01 | $\frac{1}{4} \therefore \% \% \%$ |
| F02 |  |
| F03 | $\therefore \ddot{6}$ |

Figure 3.6.9 Scatter plot of terminal F-values vs SSB for SeaStar bootstrap runs - Norwegian spring spawning herring.


Figure 3.6.10 Comparison of assessments made in 2004 and 2003 by SeaStar - Norwegian spring spawning herring.


Figure 3.6.11 Comparison of Fbar in 2003 and 2004 assessments using SeaStar - Norwegian spring spawning herring.


Figure 3.6.12 Comparison of number at age 0 in 2003 and 2004 assessments using SeaStar - Norwegian spring spawning herring.

### 4.1 Regulation of the Barents Sea Capelin Fishery

Since 1979, the Barents Sea capelin fishery has been regulated by a bilateral fishery management agreement between Russia (former USSR) and Norway. A TAC has been set separately for the winter fishery and for the autumn fishery. In recent years no autumn fishery has taken place, except for a small Russian experimental fishery. The fishery was closed from 1 May to 15 August until 1984. During the period 1984 to 1986, the fishery was closed from 1 May to 1 September. A minimum landing size of 11 cm has been in force for several years. From the autumn of 1986 to the winter of 1991, and from the autumn 1993 to the winter 1999 no fishery took place. The fishery was re-opened in the winter season 1991 and again in the winter season 1999, on a recovered stock.

In its autumn meeting of 2003, ACFM considered a harvest control rule, which was consistent with the precautionary approach. This rule defined the harvest level based on a maximum probability of $5 \%$ that SSB would fall below $\mathbf{B}_{\mathrm{lim}}$ of 200000 t (corresponding to no catch of pre-spawning capelin in 2003). ACFM also recommended that this harvest control rule be applied in 2003. (See also paragraph 4.5). During its Autumn 2003 meeting the Mixed Russian Norwegian Fishery Commission decided that no fishing should take place on Barents Sea capelin for the winter season 2004.

In 2002, the Mixed Russian Norwegian Fishery Commission agreed to adopt a management strategy based on the rule that, with $95 \%$ probability, at least 200000 t of capelin should be allowed to spawn.

### 4.2 Catch Statistics

The international catch by country and season in the years $1965-2003$ is given in Table 4.2.1. The catch by age and length groups during the spring season 2003 is given in Table 4.2.2. The total catch in winter 2003 given in Table 4.2.1 was 282000 t . This is 28000 t below the quota set for 2003. No catches were taken during autumn 2003.

### 4.3 Stock Size Estimates

### 4.3.1 Larval and 0-group estimates in 2003

Norwegian larval surveys based on Gulf III plankton samples have been carried out in June each year since 1981. The estimated total number of larvae is shown in Table 4.3.1.1. These larval abundance estimates do not show a high correlation with year class strength at age one, but should reflect the amount of larvae produced each year (Gundersen and Gjøsæter, 1998). The year 1986 was exceptional, in that no larvae were found. This may have been due to late spawning that year, and eggs may have hatched after the survey was carried out. Also in other years some spawning is known to have taken place during the summer, and offspring from such late spawning is not reflected in the larval abundance estimates in Table 4.3.1.1. Since 1997, permission has not been granted to enter the Russian EEZ during the larval survey or permission has been granted so late that it could not be employed to good purpose, and consequently the total larval distribution area has not been covered. The estimate of $11.9 \cdot 10^{12}$ larvae in 2003 is half of the estimate in 2002, at the same level as that obtained in 2001 and above the average for the period 1981-2003. During the international 0 -group surveys in August an area-based index for the abundance of 0 -group capelin is calculated (Table 4.3.1.1). Gundersen and Gjøsæter (1998) found these indices to be well correlated ( $\mathrm{r}^{2}=0.75$ ) with the 1-group acoustic estimates for the same year class obtained by the annual capelin acoustic surveys in autumn. Data points up to 1994 were included in this analysis. When this regression is updated with the survey results from 1981-2003 the parameters in the regression were changed slightly and the $\mathrm{r}^{2}$ was reduced to 0.65 . Based on this regression, (ln 1-group estimate $=$ $-1.74+1.18 \cdot \ln 0$-group index), the 0 -group index obtained in 2003 of 630 would correspond to a year class strength of 347 billion one-year-olds in autumn 2004. A year class of this size would be about 1.6 times an average year class in the period 1972-2002.

### 4.3.2 Acoustic stock size estimates in 2003

Two Russian and three Norwegian vessels jointly carried out the 2003 acoustic survey as part of an ecosystem-survey during autumn (WD by Bogstad et al.). The coverage of the total stock was considered complete. The results from the survey are given in Table 4.3.2.1, and are compared to previous years' results in Table 4.3.2.2. The stock size was estimated at 0.5 million tonnes. About $50 \%$ ( 0.28 mill t) of the stock biomass consisted of maturing fish ( $>14 \mathrm{~cm}$ ).

### 4.3.3

Other surveys
During the period of 20/02-08/03, two Russian research vessels surveyed and good coverage the area of capelins spawning stock distribution. It was about 99 thousand square nautical miles and located in the Norwegian, Russian and Grey Zones. Capelin as very scattered concentrations were recorded in the area from the $32^{\circ}-37^{\circ}$ everywhere. Typical "moving schools" practically were not found. The most part of fish was dispersed in the layers close to the bottom. It was weak migration $\mathbf{F}_{\text {low }}$ in the central part and there was not a considerable $\mathbf{F}_{\text {low }}$ in the east of the Barents Sea. The maturing stock, determined as the part of the stock exceeding 14 cm length, was estimated at $\mathbf{5 2 2 4 . 4 2 \times 1 0}{ }^{\mathbf{6}}$ ind., and $104.95 \times 10^{\mathbf{3}}$ tonnes. Capelin at the age of 3 and 4 (poor year-classes of 2001 and 2000) made up the bulk of the capelin spawning stock, in 28/60 proportion, in spring 2004.

During the Norwegian demersal fish survey in February 2004 observations of capelin by acoustics and by pelagic and demersal trawls were made. Although no stock size estimate was attempted, due to inadequate coverage and low number of pelagic trawl hauls for identification and sampling purposes, the overall impression was very low abundance of capelin found pelagically and sporadic catches of capelin in the bottom trawl hauls. Samples of cod stomachs during this period give valuable information for the modelling of maturing capelin as prey for cod (Bogstad and Gjøsæter, 2001).

### 4.4 Historical stock development

An overview of the development of the Barents Sea capelin stock in the period 1991-2003 is given in Tables 4.4.14.4.7. The methods and assumptions used for constructing the tables are explained in Appendix A to ICES 1995 Assess: 9. In that report, the complete time series back to 1973 can also be found. It should be noted that several of the assumptions and parameter values used in constructing these tables differ from those used in the assessment. For instance, in the assessment model the M-values for immature capelin are calculated using new estimates of the length at maturity and M -values for mature capelin are calculated taking the predation by cod into account. This will also affect the estimates of spawning stock biomass given in the stock summary table (Table 4.4.7). It should be noted that these values, coming from a deterministic model cannot directly be compared to those coming from the probabilistic assessment model (Bifrost) used for this stock. However, as a crude overview of the development of the Barents Sea capelin stock the tables may be adequate.

Estimates of stock in number by age group and total biomass for the period are shown in Table 4.4.1. Catch in numbers at age and total landings are shown for the spring and autumn seasons in Tables 4.4.2 and 4.4.3. Natural mortality coefficients by age group for immature and mature capelin are shown in Table 4.4.4. Stock size at 1 January in numbers at age and total biomass is shown in Table 4.4.5. Spawning stock biomass per age group is shown in Table 4.4.6. Table 4.4.7 gives an aggregated summary for the entire period 1973-2003.

### 4.5 Stock assessment autumn 2003

As decided by the Northern Pelagic and Blue Whiting Fisheries Working Group at its 2003 meeting (ICES 2003/ACFM:23), the assessment of Barents Sea capelin was left to the parties responsible for the autumn survey, i.e. IMR in Bergen and PINRO in Murmansk, who reported directly to ACFM before its autumn 2003 meeting (Bogstad et al., WD).

A probabilistic projection of the spawning stock to the time of spawning at 1 April 2004 was presented, using the spreadsheet model CapTool (implemented in the @RISK add-on for EXCEL). The projection was based on a probabilistic maturation model with parameters estimated by the model Bifrost (Gjøsæter et al. 2002) with uncertainty taken into account and data on size and composition of the cod stock (from the Arctic Fisheries Working Group, ICES 2003/ACFM:22, but made probabilistic in CapTool in accordance with the risk analysis made by the Arctic Fisheries Working Group). It was very good relationship between results of calculation SSB which geting by CapTool model and Russian winter capelin survey.

There is clearly a need for a target biomass reference point for capelin. Calculations of $\mathrm{B}_{\text {target }}$ were attempted, but were not presented because the results were considered preliminary. A $\mathbf{B}_{\mathrm{lim}}\left(\mathrm{SSB}_{\mathrm{lim}}\right)$ management approach was suggested for this stock. In 2002, the Mixed Russian Norwegian Fishery Commission agreed to adopt a management strategy based on the rule that, with $95 \%$ probability, at least 200000 t of capelin should be allowed to spawn. Consequently, 200000 t was used as a $\mathbf{B}_{\text {lim }}$.

Probabilistic prognoses for the maturing stock from October 12003 until April 12004 were made, with a CV of 0.20 on the abundance estimate. The meeting concluded that capelin recruitment in 2004 could be seriously negatively affected by the stock of young herring now found in the Barents Sea.

ACFM at its autumn 2003 meeting (ICES 2003/CRR:261) took all the points in the report into account. ACFM advised that no fishing should take place in spring 2004. This was based on adopting the forecast of the SSB using the limit reference points referred above, and following the harvest control rule that the SSB should fall below $\mathbf{B}_{\text {lim }}$ with a maximum $5 \%$ probability. ACFM also considered that adjustments of the harvest control rule should be further investigated for the purpose of taking better account of the uncertainty in the predicted estimate of spawner abundance, the likely interactions with herring, and the role of capelin as prey.

### 4.6 Management considerations

Since the assessment of the stock is directly based on the acoustic survey conducted annually in September-October, and the main fishing season does not begin until January, advice for this stock must be given during the autumn ACFM meeting and the TAC must be set by the Mixed Norwegian-Russian Fishery Commission during its meeting in November-December. As previously decided by the Northern Pelagic and Blue Whiting Fisheries Working Group, the assessment of Barents Sea capelin is left to the parties responsible for the autumn survey, i.e. IMR in Bergen and PINRO in Murmansk, who will meet in Kirkenes from 05-08 October 2004 and report directly to the 2004 ACFM autumn meeting.

### 4.7 Sampling

The sampling from scientific surveys and from commercial fishing on capelin in 2003 and winter 2004 is summarised below:

| Investigation | No. of samples | Length measurements | Aged individuals |
| :--- | :---: | :---: | :---: |
| Russian capelin investigation winter 2003 | 135 | 4808 | - |
| Russian fishery winter-spring 2003 | 114 | 19449 | 800 |
| Norwegian capelin investigations winter 2003 | 163 | 5581 | 1742 |
| Norwegian fishery winter-spring 2003 | 44 | 4340 | 1461 |
| Acoustic survey autumn 2003 (Norway) | 196 | 7707 | 2231 |
| Acoustic survey autumn 2003 (Russia) | 142 | 7330 | 725 |
| Other samples 2003 (Norway) | 139 | 8500 | - |
| Other sample 2003 (Russia) | 79 | 5739 | 50 |
| Norwegian capelin investigations winter 2004 | 213 | 6590 | 1994 |
| Russian capelin investigation winter 2004 | 167 | 9368 | 883 |

Table 4.2.1 Barents Sea CAPELIN. International catch (' 000 t ) as used by the Working Group.

| Year | Winter |  |  |  | Summer-Autumn |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Norway | Russia | Others | Total | Norway | Russia | Total |  |
| 1965 | 217 | 7 | 0 | 224 | 0 | 0 | 0 | 224 |
| 1966 | 380 | 9 | 0 | 389 | 0 | 0 | 0 | 389 |
| 1967 | 403 | 6 | 0 | 409 | 0 | 0 | 0 | 409 |
| 1968 | 460 | 15 | 0 | 475 | 62 | 0 | 62 | 537 |
| 1969 | 436 | 1 | 0 | 437 | 243 | 0 | 243 | 680 |
| 1970 | 955 | 8 | 0 | 963 | 346 | 5 | 351 | 1314 |
| 1971 | 1300 | 14 | 0 | 1314 | 71 | 7 | 78 | 1392 |
| 1972 | 1208 | 24 | 0 | 1232 | 347 | 11 | 358 | 1591 |
| 1973 | 1078 | 35 | 0 | 1112 | 213 | 10 | 223 | 1336 |
| 1974 | 749 | 80 | 0 | 829 | 237 | 82 | 319 | 1149 |
| 1975 | 559 | 301 | 43 | 903 | 407 | 129 | 536 | 1439 |
| 1976 | 1252 | 231 | 0 | 1482 | 739 | 366 | 1105 | 2587 |
| 1977 | 1441 | 345 | 2 | 1788 | 722 | 477 | 1199 | 2987 |
| 1978 | 784 | 436 | 25 | 1245 | 360 | 311 | 671 | 1916 |
| 1979 | 539 | 343 | 5 | 887 | 570 | 326 | 896 | 1783 |
| 1980 | 539 | 253 | 9 | 801 | 459 | 388 | 847 | 1648 |
| 1981 | 784 | 428 | 28 | 1240 | 454 | 292 | 746 | 1986 |
| 1982 | 568 | 260 | 5 | 833 | 591 | 336 | 927 | 1760 |
| 1983 | 751 | 374 | 36 | 1161 | 758 | 439 | 1197 | 2358 |
| 1984 | 330 | 257 | 42 | 628 | 481 | 367 | 849 | 1477 |
| 1985 | 340 | 234 | 17 | 590 | 113 | 164 | 278 | 868 |
| 1986 | 72 | 51 | 0 | 123 | 0 | 0 | 0 | 123 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 528 | 156 | 20 | 704 | 31 | 195 | 226 | 929 |
| 1992 | 620 | 247 | 24 | 891 | 73 | 159 | 232 | 1123 |
| 1993 | 402 | 170 | 14 | 586 | 0 | 0 | 0 | 586 |
| 1994 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 |
| 1998 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 |
| 1999 | 46 | 32 | 0 | 78 | 0 | 23 | 23 | 101 |
| 2000 | 283 | 95 | 8 | 386 | 0 | 28 | 28 | 414 |
| 2001 | 368 | 180 | 8 | 557 | 0 | 11 | 11 | 568 |
| 2002 | 391 | 228 | 17 | 635 | 0 | 16 | 16 | 651 |
| 2003 | 190 | 93 | 0 | 282 | 0 | 0 | 0 | 282 |

Table 4.2.2 Barents Sea CAPELIN. International catch in number $\left(10^{6}\right)$ and biomass $(\mathrm{t})$ during the spring season 2003, as used by the Working Group

| Length cm | Age 1 |  | Age 2 |  | Age 3 |  | Age 4 |  | Age 5+ |  | Sum |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | N | B | N | B | N | B | N | B | N | B | N | \% | B | \% |
| 5.0-5.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5.5-6.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6.0-6.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6.5-7.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7.0-7.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7.5-8.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8.0-8.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8.5-9.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9.0-9.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9.5-10.0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 10.0- | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 10.5- | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11.0- | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11.5- | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 12.0- | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 12.5- | 0 | 0 | 0 | 0 | 4480 | 41 | 0 | 0 | 0 | 0 | 4480 | 0 | 41 | 0 |
| 13.0- | 0 | 0 | 0 | 0 | 227047 | 2138 | 4692 | 43 | 0 | 0 | 231739 | 2 | 2180 | 1 |
| 13.5- | 0 | 0 | 0 | 0 | 248431 | 2598 | 123197 | 1262 | 0 | 0 | 371627 | 3 | 3860 | 1 |
| 14.0- | 0 | 0 | 0 | 0 | 462853 | 5347 | 330091 | 3869 | 0 | 0 | 792944 | 6 | 9217 | 3 |
| 14.5- | 0 | 0 | 0 | 0 | 311292 | 4088 | 776663 | 10559 | 4692 | 72 | 1092647 | 8 | 14718 | 5 |
| 15.0- | 0 | 0 | 0 | 0 | 352191 | 5178 | 1058475 | 15343 | 0 | 0 | 1410666 | 11 | 20521 | 7 |
| 15.5- | 0 | 0 | 0 | 0 | 168810 | 2828 | 1364395 | 23405 | 36993 | 678 | 1570199 | 12 | 26910 | 10 |
| 16.0- | 0 | 0 | 0 | 0 | 139504 | 2736 | 1482273 | 28883 | 53564 | 1091 | 1675342 | 13 | 32710 | 12 |
| 16.5- | 0 | 0 | 0 | 0 | 132094 | 2741 | 1534349 | 34363 | 56778 | 1409 | 1723222 | 13 | 38513 | 14 |
| 17.0- | 0 | 0 | 0 | 0 | 21942 | 581 | 1370535 | 35220 | 36723 | 777 | 1429201 | 11 | 36579 | 13 |
| 17.5- | 0 | 0 | 0 | 0 | 36700 | 1094 | 1188436 | 34896 | 78429 | 2270 | 1303564 | 10 | 38259 | 14 |
| 18.0- | 0 | 0 | 0 | 0 | 9384 | 302 | 759772 | 24561 | 95860 | 3037 | 865016 | 6 | 27900 | 10 |
| 18.5- | 0 | 0 | 0 | 0 | 33854 | 1096 | 454478 | 16325 | 46960 | 1661 | 535292 | 4 | 19082 | 7 |
| 19.0- | 0 | 0 | 0 | 0 | 0 | 0 | 235440 | 8719 | 4692 | 175 | 240132 | 2 | 8895 | 3 |
| 19.5- | 0 | 0 | 0 | 0 | 0 | 0 | 52999 | 2207 | 0 | 0 | 52999 | 0 | 2207 | 1 |
| 20.0- | 0 | 0 | 0 | 0 | 0 | 0 | 16624 | 790 | 0 | 0 | 16624 | 0 | 790 | 0 |
| 20.5- | 0 | 0 | 0 | 0 | 0 | 0 | 372 | 20 | 0 | 0 | 372 | 0 | 20 | 0 |
| 21.0- | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 21.5- | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sum | 0 | 0 | 0 | 0 | 2148582 | 30766 | 1075279 | 240464 | 414691 | 11170 | 1331606 | 100 | 282400 | 100 |

Table 4.3.1.1 Barents Sea CAPELIN. Larval abundance estimate $\left(10^{12}\right)$ in June, and 0 -group index in August.

| Year | Larval <br> abundance | 0-group <br> index |
| :---: | :---: | :---: |
| 1981 | 9.7 | 570 |
| 1982 | 9.9 | 393 |
| 1983 | 9.9 | 589 |
| 1984 | 8.2 | 320 |
| 1985 | 8.6 | 110 |
| 1986 | 0.0 | 125 |
| 1987 | 0.3 | 55 |
| 1988 | 0.3 | 187 |
| 1989 | 7.3 | 1300 |
| 1990 | 13.0 | 324 |
| 1991 | 3.0 | 241 |
| 1992 | 7.3 | 26 |
| 1993 | 3.3 | 43 |
| 1994 | 0.1 | 58 |
| 1995 | 0.0 | 43 |
| 1996 | 2.4 | 291 |
| 1997 | 6.9 | 522 |
| 1998 | 14.1 | 428 |
| 1999 | 36.5 | 722 |
| 2000 | 19.1 | 303 |
| 2001 | 10.7 | 221 |
| 2002 | 22.4 | 327 |
| 2003 | 11.9 | 630 |
|  |  |  |

Table 4.3.2.1 Barents Sea CAPELIN. Estimated stock size from the acoustic survey in September-October 2003. Based on TS value $19.1 \log \mathrm{~L}-74.0 \mathrm{~dB}$, corresponding to $\sigma=5.0 \cdot 10^{7} \cdot \mathrm{~L}^{1.91}$.


Based on TS value: $19.1 \log \mathrm{~L}-74.0$, corresponding to $\sigma=5.0 \cdot 10^{-7} \cdot \mathrm{~L}^{1.9}$

Table 4.3.2.2 Barents Sea CAPELIN. Stock size in numbers by age, total stock biomass and biomass of the maturing component. Stock in numbers (unit: $10^{9}$ ) and stock and maturing stock biomass (unit: $10^{3}$ tonnes) are given at 1 . October.

| Year | $\text { Stock in numbers }\left(10^{9}\right)$ |  |  |  |  |  | Stock in weight ('000 t) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |
|  | $\text { Age } 1$ | $\text { Age } 2$ | Age 3 | Age 4 | Age 5 | Total | Total | Maturing |
| 1973 | 528 | 375 | 40 | 17 | 0 | 961 | 5144 | 1350 |
| 1974 | 305 | 547 | 173 | 3 | 0 | 1029 | 5733 | 907 |
| 1975 | 190 | 348 | 296 | 86 | 0 | 921 | 7806 | 2916 |
| 1976 | 211 | 233 | 163 | 77 | 12 | 696 | 6417 | 3200 |
| 1977 | 360 | 175 | 99 | 40 | 7 | 681 | 4796 | 2676 |
| 1978 | 84 | 392 | 76 | 9 | 1 | 561 | 4247 | 1402 |
| 1979 | 12 | 333 | 114 | 5 | 0 | 464 | 4162 | 1227 |
| 1980 | 270 | 196 | 155 | 33 | 0 | 654 | 6715 | 3913 |
| 1981 | 403 | 195 | 48 | 14 | 0 | 660 | 3895 | 1551 |
| 1982 | 528 | 148 | 57 | 2 | 0 | 735 | 3779 | 1591 |
| 1983 | 515 | 200 | 38 | 0 | 0 | 754 | 4230 | 1329 |
| 1984 | 155 | 187 | 48 | 3 | 0 | 393 | 2964 | 1208 |
| 1985 | 39 | 48 | 21 | 1 | 0 | 109 | 860 | 285 |
| 1986 | 6 | 5 | 3 | 0 | 0 | 14 | 120 | 65 |
| 1987 | 38 | 2 | 0 | 0 | 0 | 39 | 101 | 17 |
| 1988 | 21 | 29 | 0 | 0 | 0 | 50 | 428 | 200 |
| 1989 | 189 | 18 | 3 | 0 | 0 | 209 | 864 | 175 |
| 1990 | 700 | 178 | 16 | 0 | 0 | 894 | 5831 | 2617 |
| 1991 | 402 | 580 | 33 | 1 | 0 | 1016 | 7287 | 2248 |
| 1992 | 351 | 196 | 129 | 1 | 0 | 678 | 5150 | 2228 |
| 1993 | 2 | 53 | 17 | 2 | 2 | 75 | 796 | 330 |
| 1994 | 20 | 3 | 4 | 0 | 0 | 28 | 200 | 94 |
| 1995 | 7 | 8 | 2 | 0 | 0 | 17 | 193 | 118 |
| 1996 | 82 | 12 | 2 | 0 | 0 | 96 | 503 | 248 |
| 1997 | 99 | 39 | 2 | 0 | 0 | 140 | 911 | 312 |
| 1998 | 179 | 73 | 11 | 1 | 0 | 263 | 2056 | 931 |
| 1999 | 156 | 101 | 27 | 1 | 0 | 285 | 2776 | 1718 |
| 2000 | 449 | 111 | 34 | 1 | 0 | 595 | 4273 | 2099 |
| 2001 | 114 | 219 | 31 | 1 | 0 | 364 | 3630 | 2019 |
| 2002 | 60 | 91 | 50 | 1 | 0 | 201 | 2210 | 1290 |
| 2003 | 82 | 10 | 11 | 1 | 0 | 104 | 533 | 280 |

Table 4.4.1 Barents Sea CAPELIN. Estimated stock size in numbers (unit: $10^{9}$ ) by age group and total, and biomass (' 000 t ) of total stock, by 1. August, back-calculated from the survey in September-October.

| Age | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 29.5 | 8.3 | 88.9 | 111.8 | 188.4 | 171.4 | 474.7 | 128.0 | 67.3 | 92.9 |
| 2 | 5.1 | 9.4 | 12.5 | 44.2 | 76.5 | 111.5 | 116.8 | 246.6 | 102.3 | 13.0 |
| 3 | 6.4 | 1.6 | 2.2 | 2.2 | 12.1 | 27.9 | 35.9 | 33.0 | 54.4 | 14.6 |
| 4 | 0.3 | 0.4 | 0.1 | 0.1 | 0.7 | 0.9 | 0.8 | 1.2 | 0.6 | 1.9 |
| 5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 |
| Sum | 41.4 | 19.7 | 103.7 | 158.3 | 277.8 | 311.7 | 628.4 | 408.8 | 224.7 | 122.4 |
| Biomass | 259 | 189 | 467 | 866 | 1860 | 2580 | 3840 | 3480 | 2122 | 662 |

Table 4.4.2 Barents Sea CAPELIN. Catch in numbers (unit: $10^{9}$ ) by age group and total landings (' 000 t ) in the spring season.

| Age | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 |
| 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.5 | 0.4 | 0.0 |
| 3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 1.6 | 5.5 | 7.6 | 10.0 | 2.1 |
| 4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 1.2 | 8.4 | 12.1 | 14.2 | 10.8 |
| 5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 1.0 | 2.2 | 0.7 | 1.4 |
| Sum | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 3.0 | 15.1 | 22.5 | 25.3 | 14.3 |
| Landings | 0 | 0 | 0 | 0 | 0 | 78 | 386 | 557 | 635 | 635 |

Table 4.4.3 Barents Sea CAPELIN. Catch in numbers (unit: $10^{9}$ ) by age group and total landings (' 000 t ) in the autumn season.

| Age | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.4 | 0.1 | 0.0 | 0.0 | 0.0 |
| 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.9 | 0.9 | 0.4 | 0.3 | 0.0 |
| 3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.4 | 0.2 | 0.6 | 0.0 |
| 4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Sum | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 1.6 | 1.5 | 0.6 | 0.9 | 0.0 |
| Landings | 0 | 0 | 0 | 1 | 1 | 23 | 28 | 11 | 16 | 0 |

Table 4.4.4 Barents Sea CAPELIN. Natural mortality coefficients (per month) for immature fish $\left(\mathrm{M}_{\mathrm{imm}}\right)$, used for the whole year, and for mature fish (per season) ( $\mathrm{M}_{\mathrm{mat}}$ ) used January to March, by age group and average for age groups 1-5.

|  | 1994 |  | 1995 |  | 1996 |  | 1997 |  | 1998 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | $\mathrm{M}_{\text {imm }}$ | $\mathrm{M}_{\text {mat }}$ | $\mathrm{M}_{\text {imm }}$ | $\mathrm{M}_{\text {mat }}$ | $\mathrm{M}_{\mathrm{imm}}$ | $\mathrm{M}_{\text {mat }}$ | $\mathrm{M}_{\text {imm }}$ | $\mathrm{M}_{\text {mat }}$ | $\mathrm{M}_{\text {imm }}$ | $\mathrm{M}_{\text {mat }}$ |
| 1 | 0.201 | 0.602 | 0.073 | 0.219 | 0.041 | 0.122 | 0.062 | 0.185 | 0.026 | 0.077 |
| 2 | 0.201 | 0.602 | 0.073 | 0.219 | 0.041 | 0.122 | 0.062 | 0.185 | 0.026 | 0.077 |
| 3 | 0.201 | 0.602 | 0.019 | 0.058 | 0.041 | 0.122 | 0.062 | 0.185 | 0.071 | 0.212 |
| 4 | 0.282 | 0.847 | 0.044 | 0.133 | 0.050 | 0.149 | 0.014 | 0.041 | 0.071 | 0.212 |
| 5 | 0.282 | 0.847 | 0.044 | 0.133 | 0.050 | 0.149 | 0.014 | 0.041 | 0.071 | 0.212 |
| Avr | 0.221 | 0.700 | 0.052 | 0.152 | 0.043 | 0.133 | 0.042 | 0.127 | 0.053 | 0.158 |

Table 4.4.4 (Continued)

|  | 1999 |  | 2000 |  | 2001 |  | 2002 |  | 2003 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | $\mathrm{M}_{\text {imm }}$ | $\mathrm{M}_{\text {mat }}$ | $\mathrm{M}_{\text {imm }}$ | $\mathrm{M}_{\text {mat }}$ | $\mathrm{M}_{\text {imm }}$ | $\mathrm{M}_{\text {mat }}$ | $\mathrm{M}_{\text {imm }}$ | $\mathrm{M}_{\text {mat }}$ | $\mathrm{M}_{\text {imm }}$ | $\mathrm{M}_{\text {mat }}$ |
| 1 | 0.047 | 0.142 | 0.028 | 0.083 | 0.060 | 0.180 | 0.060 | 0.180 | 0.060 | 0.180 |
| 2 | 0.047 | 0.142 | 0.028 | 0.083 | 0.060 | 0.180 | 0.060 | 0.180 | 0.152 | 0.456 |
| 3 | 0.025 | 0.074 | 0.026 | 0.079 | 0.040 | 0.120 | 0.040 | 0.120 | 0.142 | 0.426 |
| 4 | 0.025 | 0.074 | 0.026 | 0.079 | 0.040 | 0.120 | 0.040 | 0.120 | 0.142 | 0.426 |
| 5 | 0.025 | 0.074 | 0.026 | 0.079 | 0.040 | 0.120 | 0.040 | 0.120 | 0.142 | 0.426 |
| Avr | 0.034 | 0.101 | 0.027 | 0.080 | 0.048 | 0.144 | 0.048 | 0.144 | 0.128 | 0.383 |

Table 4.4.5 Barents Sea CAPELIN. Estimated stock size in numbers (unit: $10^{9}$ ) by age group and total, and biomass ('000 t) of total stock, by 1. January.

| Age | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 120.3 | 13.8 | 118.2 | 172.0 | 225.5 | 238.5 | 576.1 | 194.7 | 102.3 | 141.3 |
| 2 | 1.4 | 10.8 | 5.7 | 72.5 | 82.2 | 165.8 | 135.3 | 413.3 | 94.6 | 75.4 |
| 3 | 33.3 | 1.9 | 6.5 | 10.2 | 32.5 | 67.3 | 88.1 | 100.9 | 182.6 | 44.5 |
| 4 | 9.8 | 2.4 | 1.4 | 1.8 | 1.6 | 8.5 | 24.7 | 31.1 | 27.0 | 0.5 |
| 5 | 1.3 | 0.1 | 0.3 | 0.1 | 0.1 | 0.5 | 0.8 | 0.7 | 0.9 | 0.0 |
| Sum | 166.1 | 28.9 | 132.2 | 256.6 | 341.9 | 480.6 | 824.9 | 740.6 | 407.5 | 261.7 |
| Biomass | 737 | 156 | 313 | 779 | 1240 | 2456 | 3571 | 4558 | 3539 | 2008 |

Table 4.4.6 Barents Sea CAPELIN. Estimated spawning stock biomass (‘000 t) by 1. April.

| Age | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 1 | 3 | 1 | 1 | 2 | 24 | 5 | 0 | 169 |
| 3 | 34 | 15 | 71 | 175 | 217 | 650 | 819 | 943 | 867 | 663 |
| 4 | 60 | 38 | 24 | 49 | 34 | 193 | 472 | 539 | 339 | 0 |
| 5 | 11 | 1 | 7 | 2 | 2 | 10 | 0 | 0 | 9 | 0 |
| Sum | 105 | 55 | 105 | 228 | 254 | 856 | 1315 | 1487 | 1215 | 832 |

Table 4.4.7 Barents Sea CAPELIN. Stock summary table. Recruitment (number of 1 year old fish (unit: $10^{9}$ ) and stock biomass (' 000 t ) given at 1 . August, spawning stock (' 000 t ) at time of spawning (1. April). Landings (' 000 t ) are the sum of the total landings in the two fishing seasons within the year indicated. The SSB is obtained by projecting the stock forward assuming a natural mortality that does not take the current predation mortality fully into account.

| Year | Stock biomass | Recruitment Age 1 | Spawning stock biomass | Landings |
| :---: | :---: | :---: | :---: | :---: |
| 1965 |  |  |  | 224 |
| 1966 |  |  |  | 389 |
| 1967 |  |  |  | 409 |
| 1968 |  |  |  | 537 |
| 1969 |  |  |  | 680 |
| 1970 |  |  |  | 1314 |
| 1971 |  |  |  | 1392 |
| 1972 | 5831 |  |  | 1592 |
| 1973 | 6630 | 1140 | 1242 | 1336 |
| 1974 | 7121 | 737 | 343 | 1149 |
| 1975 | 8841 | 494 | 90 | 1439 |
| 1976 | 7584 | 433 | 1147 | 2587 |
| 1977 | 6254 | 830 | 890 | 2987 |
| 1978 | 6119 | 855 | 460 | 1916 |
| 1979 | 6576 | 551 | 193 | 1783 |
| 1980 | 8219 | 592 | 87 | 1648 |
| 1981 | 4489 | 466 | 1731 | 1986 |
| 1982 | 4205 | 611 | 546 | 1760 |
| 1983 | 4772 | 612 | 47 | 2358 |
| 1984 | 3303 | 183 | 171 | 1477 |
| 1985 | 1087 | 47 | 106 | 868 |
| 1986 | 157 | 9 | 13 | 123 |
| 1987 | 107 | 46 | 16 | 0 |
| 1988 | 361 | 22 | 11 | 0 |
| 1989 | 771 | 195 | 141 | 0 |
| 1990 | 4901 | 708 | 179 | 0 |
| 1991 | 6647 | 415 | 1584 | 929 |
| 1992 | 5371 | 396 | 998 | 1123 |
| 1993 | 991 | 3 | 460 | 586 |
| 1994 | 259 | 30 | 105 | 0 |
| 1995 | 189 | 8 | 55 | 0 |
| 1996 | 467 | 89 | 105 | 0 |
| 1997 | 866 | 112 | 228 | 1 |
| 1998 | 1860 | 188 | 254 | 1 |
| 1999 | 2580 | 171 | 856 | 106 |
| 2000 | 3840 | 475 | 1315 | 414 |
| 2001 | 3480 | 128 | 1487 | 568 |
| 2002 | 2122 | 67 | 1215 | 651 |
| 2003 | 662 | 93 | 832 | 282 |

### 5.1 The Fishery

### 5.1.1 Regulation of the fishery

The fishery is based on maturing capelin, i.e. that part of each year class which spawns at age 3 as well as those fish at age 4 , that did not mature and spawn at age 3 . The abundance of the immature component is difficult to assess before their recruitment to the adult stock at ages 2 and 3 . This is especially true of the age 3 immatures.

The fishery of the Icelandic capelin has, therefore, been regulated by preliminary catch quotas set prior to each fishing season (July-March). Predictions of TACs have been computed from autumn survey data the year before on the abundance of 1 and 2 year old capelin. The process includes historical relationships between such data and the backcalculated abundance of the same year classes, growth rate and stock in numbers, natural mortality and the provision of a remaining spawning stock biomass of 400000 t . Final catch quotas for each season have then been set according to the results of acoustic surveys of the maturing, fishable stock, carried out in autumn (October-November) and/or winter (January/February) in that fishing season. A more detailed description of the method is given in Section 1.3.5. A summary of the results of this catch regulation procedure is given in Table 5.1.1.

Over the years the fishery has been closed during April - late June and the season has started in July/August or later, depending on the state of the stock. Due to very low stock abundance there was a fishing ban lasting from December 1981 to November 1983. In addition, areas with high abundances of juvenile age 1 and 2 capelin (in the shelf region off NW-, N - and NE-Iceland) have usually been closed to the summer and autumn fishery.

### 5.1.2 The fishery in the 2003/2004 season

In accordance with a previously determined procedure, ACFM recommended that the preliminary TAC should not exceed 555000 t . This is $2 / 3$ of the total TAC prediction of 835000 t for the $2003 / 2004$ season. This advice was accepted by all parties concerned.

The season opened on 20 June, with insignificant catches taken off the northeast and eastern north coast of Iceland and later more intensive fishery developed in deep waters off the shelf edge north of Iceland. The fishing grounds soon shifted to the northwest and north. Catch rates were low in the beginning but improved in July as the capelin migrated northward. Towards the end of July, the northward migration stopped and the capelin soon scattered. At that time the total catch was about 200000 t , mainly taken in July on both sides of the Iceland/Greenland EEZ between about $68^{\circ} \mathrm{N}$ and $69^{\circ} 30^{\prime} \mathrm{N}$. After July the capelin remained scattered and no catches were made for the rest of the year. The small fishery, which in later years has been conducted near the shelfbreak off the eastern north coast of Iceland in December, did not materialize.

The total catch in the 2003 summer and autumn season was just under 200000 t (Table 5.2.1).

In the first half January 2004 the research ship and a large fleet of scouting vessels only found scattered concentrations of capelin. In view of the uncertain stock prognosis (see section 5.7) a complete fishing ban was enforced, to last until a sufficiently large assessment of fishable stock abundance had been obtained. In the meantime, the Marine Research Institute and the Association of Owners of Fishing Boats arranged for a watch by commercial vessels of the most likely fishing areas in the near future.

On 17 January, one of these fishing vessels found large fishable concentrations of adult capelin at the shelf edge off the eastern north coast of Iceland. These, as well as less dense concentrations further south and east, were assessed by the research vessel in the following days and the fishing ban was subsequently lifted.

During the last week of January and in early February, the fishery gradually moved south, generally following the shelf edge east of Iceland. Catch rates remained high during most of this period. For this reason the large concentrations, which were located by the Icelandic research vessel about 40-60 nautical miles farther east, were not fished at the time.

The first spawning run entered the warm Atlantic water off the southeast coast round 20 February, which is unusually late. From there, the capelin migrated slowly westward in near-shore areas. As usual, catch rates were high in the shallow water area until early March, when stormy weather prevented fishing for a week. As often happens under such circumstances, spawning began and the capelin did not form fishable schools again in but a few places. Due to the late arrival and slow migration speed west along the south coast, only a few schools rounded the Reykjanes promontory to
spawn off the west coast. Some of the spawners, located in the easternmost area around mid-February, probably followed the earlier migration(s) and intermingled with them in late February/early March. The remainder was never fished and their fate is therefore unknown.

The total catch during the 2004 winter season was 543000 t (Table 5.2.1). Due to the abrupt end of the fishery before mid-March, about 135000 t of the TAC, finally set for the 2003/04 season, was not taken.

### 5.2 Catch Statistics

The total annual catch of capelin in the Iceland-East Greenland-Jan Mayen area since 1964 is given by weight, season, and fleet in Table 5.2.1.

The total catch in numbers during the summer/autumn 1982-2003 and winter 1983-2004 seasons is given by age and years in Tables 5.2.2 and 5.2.3.

The distribution of the catch during the summer-autumn 2003 and winter 2004 seasons is given by length groups at age in Tables 5.2.4 and 5.2.5.

### 5.3 Surveys of Stock Abundance

### 5.3.1 0-group surveys

The distribution and abundance of 0 -group capelin in the Iceland-East Greenland-Jan Mayen area has been recorded during surveys carried out in August since 1970. The survey methods and computations of abundance indices were described by Vilhjálmsson and Fridgeirsson (1976). The abundance indices of 0-group capelin, divided according to areas, are given in Table 5.3.1.1.

Acoustic estimates of the abundance of 1 -group capelin have also been obtained during the August 0 -group surveys (e.g. Vilhjálmsson 1994). However, due to the large variability of this time series, the August abundance indices have not been used for stock projections for more than a decade. Directed collection of data on juvenile age 1 and 2 capelin in August has been discontinued. The abundance of 1-group capelin by number, mean length and weight for 1983-2001 is given in Table 5.3.1.2.

### 5.3.2 Stock abundance in autumn 2003 and winter 2004

### 5.3.2.1 The adult fishable stock

A survey specifically aimed at estimating the adult fishable stock was not made in the autumn of 2003. Like in autumn 2002, hardly any adult capelin were located during the assessment of juvenile capelin in November 2003.

Due to the unusually large uncertainty associated with the prediction of the fishable stock size in the 2003/04 season, twelve fishing vessels carried out a coordinated pilot survey off NW- N- and E-Iceland in the first week of January 2004. Although capelin were recorded over fairly wide areas, densities were very low. On the basis of the findings by the scouting vessels, a research vessel carried out an acoustic survey of the most promising areas during 3-12 January. The total adult stock biomass came to only 127000 t and as a results all fishing permits were temporarily revoked.

Following these catastrophic results, it was arranged that the most likely areas be monitored by fishing vessels during the latter half of January in order to ensure the shortest possible time needed for the research vessel to obtain a new estimate. On 16 January and in the following few days, a large migration was located near the shelf break off the eastern north coast as well as smaller concentrations of lesser densities further to the east and southeast. All of these capelin were assessed by acoustics during 18-30 January and the fishing ban was lifted. The total abundance of capelin in the area was estimated at 780000 t , with the 2001 year class strongly dominating both by numbers and weight.

Since weather conditions were inadequate, especially when working off the northern east coast in late January, it was decided to run a third survey of the adult fishable stock east of Iceland. This survey took place during 10-18 February and may be divided into two parts. In the beginning, capelin were assessed in the fishing area at and just out from the shelf edge off the southern east coast. Dense schools were recorded in the southernmost part of this area where catch rates were also high. Densities declined rapidly further north and at about $65^{\circ} \mathrm{N}$ had become very low. However, another area containing very dense and large concentrations was discovered $40-60$ nautical miles farther east, from about
$64^{\circ} 50^{\prime} \mathrm{N}$ south to $64^{\circ} \mathrm{N}$, between $09^{\circ} \mathrm{W}$ and $10^{\circ} \mathrm{W}$. Scattered capelin were recorded between $64^{\circ} \mathrm{N}$ and $63^{\circ} 50^{\prime} \mathrm{N}, 09^{\circ}-$ $09^{\circ} 30^{\prime} \mathrm{W}$ and small schools of capelin were also recorded south of this by a Faroese research vessel in the beginning of the third week of February. Since the bulk of the capelin in this easternmost area were slowly migrating to the northwest, there is little doubt that part of the adult stock entered Faroese waters in February 2004. However, on the basis of the available evidence it cannot be determined how far south the capelin distribution in this eastern area reached. East of $10^{\circ} \mathrm{W}$, the capelin were somewhat smaller and their maturation less advanced than that of the capelin in the western area. The total abundance was estimated at about 925000 t , of which approximately $1 / 3$ was recorded in the western area and $2 / 3$ in the area east of $10^{\circ} \mathrm{W}$.

The total biomass of adult capelin, assessed north and later east of Iceland in January and February 2004 varied greatly between surveys. In retrospect it seems clear, that neither of the January estimates can have covered but parts of the total distribution area. It is obvious, that the first survey did not reach far enough to the north (stopped at $68^{\circ} \mathrm{N}$ ), while weather disturbed the latter part of the late January survey. The large migration, located at the shelf break off the easternmost north coast around mid-January, obviously entered Icelandic waters from the northwest, migrated south along and just outside the shelf edge east of Iceland and formed the bulk of the capelin that was fished east of Iceland in late January-early February. On the other hand the more oceanic concentrations appear, at least in part, to have arrived from due north, i.e. from more easterly feeding areas than the other.

Details of the January and February 2004 acoustic estimates of adult capelin are given in Tables 5.3.2.1.2-5.3.2.1.4.

### 5.3.2.2 Estimates of immature capelin

An acoustic survey was carried out in November in order to assess the abundance of immature age 1 and 2 capelin (year classes 2000 and 2001). The survey only recorded juvenile capelin off the shelf northwest of the Vestfirdir peninsula. These capelin were extremely scattered and the total estimate was only 4.6 and $3.1 * 10^{9}$ by number of the 2002 and 2001 year classes, respectively. Like in autumn 2002, hardly any adult capelin were recorded.

The winter (January-February) 2004 assessment surveys of the mature fishable stock also recorded immature capelin east of Iceland and off the eastern north coast, but their numbers were very low on all occasions.

During 13-22 April 2004 a survey of the area north and west of the Vestfirdir peninsula (NW-Iceland) in otder to locate and assess eventual concentrations of young adolescent capelin, which might be located there as the corresponding stock component was in April 2003. Survey coverage was limited to the north by drift ice at about $67^{\circ} 40^{\prime} \mathrm{N}$, whereas in the Denmark Strait the ice edge was at approximately 15 nautical miles on the Icelandic side of the equidistance line between Greenland and Iceland. The 2004 April survey recorded no capelin at all, although it covered the area west of Vestfirdir where the highest concentrations were located in April 2003.

While the above survey results may correctly reflect an extremely low abundance of the 2002 year class, there may be environmental factors at play, affecting i.e. the geographic distribution of the juvenile stock (see Section 5.7) and therefore its availability to the survey as it was run.

### 5.4 Historical Stock Abundance

The historical estimates of stock abundance are based on the "best" acoustic estimates of the abundance of maturing capelin in autumn and/or winter surveys, the "best" in each case being defined as that estimate on which the final decision of TAC was based. Taking account of the catch in number and a monthly natural mortality rate of $\mathrm{M}=0.035$ (ICES 1991/Assess:17), abundance estimates of each age group are then projected to the appropriate point in time. Since natural mortality rates of juvenile capelin are not known, their abundance by number has been projected using the same natural mortality rate.

The annual abundance by number and weight at age for mature and immature capelin in the Iceland-East Greenland-Jan Mayen area has been calculated with reference to 1 August and 1 January of the following year for the 1978/792003/04 seasons. The results are given in Tables 5.4.1 and 5.4.2 (1 August and 1 January, respectively). Table 5.4.2 also gives the remaining spawning stock by number and biomass in March/April 1979-2003.

The observed annual mean weight at age, obtained from catch and survey data from January, was used to calculate the stock biomass on 1 January. The observed average weight at age of adult capelin in autumn (Table 5.5.1.2) is used to calculate stock biomass of the maturing components in summer. Because there is a small weight increase among mature capelin in February and March, the remaining spawning stock biomass is underestimated.

### 5.5.1 Stock prognosis and TAC in the 2003/2004 season

The models (ICES 1993/Assess:6; Section 3.1.5) for predicting the numbers of maturing capelin of ages 2 and 3 from the November 2001 acoustic assessment of the 1999 and 1998 year classes gave estimates of 63.0 and $15.6 * 10^{9}$ maturing 2- and 3-group capelin on 1 August 2003.

The Working Group decided that the Aril 2003 estimate of the abundance of immature 2-group capelin (year class 2001) could be used for predicting the abundance of maturing capelin of the 2001 year class on 1 August 2003. This was done by projecting the April 2003 estimate back to November 2002, using a monthly $\mathrm{M}=0.035$, i.e. to the time of estimations used for predicting the younger year class in the fishable and spawning stock for the last two decades. On the other hand, the contribution by the 2000 year class was, as usual, from the historic relationship between the total abundance of each year class at age 3 and the contribution of the same year class at age 4 . In order to correspond to the timeline used in the past, the data for the older component (2000 year class) were projected back in time to November 2003 using a monthly $\mathrm{M}=0.035$. The resulting projections of 63.0 and 15.6 billion mature fish belonging to the 2001 and 2000 year classes respectively, are given in Table 5.5.1.1.

During the last ten years the weight at age of adult capelin has been inversely related to adult stock abundance and simple linear regressions result in $\mathrm{R}^{2}=0.64$ and 0.68 for age groups 2 and 3 , respectively. These two regression plots are shown in Figure 5.5.3.2. Applying the appropriate regression equations; $y=-0.034 x+19.4 ; r^{2}=0.64 ; p<0.05$ for the younger component, and $y=-0.068 x+29.0 ; r^{2}=0.68 ; p<0.05$ for the older one and using the predicted abundance of age groups 2 and 3 on 1 August 2003 combined, i.e. $78.6 * 10^{9}$ fish, results in estimated mean weights of 16.7 and 23.9 g for age groups 2 and 3 respectively.

Applying the estimated mean weight, results in a predicted TAC of 845000 t spread evenly from August 2003-March 2004. This corresponds to a preliminary TAC of 555000 t . As in previous years, decisions on the final TAC for the 2003/2004 season should be based on surveys carried out in October/November 2003 and January/February 2004.

The fishable stock biomass, obtained by multiplying the stock in numbers by the predicted mean weight of maturing capelin in autumn, was projected forward to spawning time in March 2003 assuming a monthly $\mathrm{M}=0.035$ and a remaining spawning stock of 400000 t . This resulted in a predicted TAC of 845000 t spread evenly over August 2003March 2004 (Table 5.5.1.3). Using the same approach as in previous years, i.e. that the preliminary TAC be set at $2 / 3$ of the predicted total for the season, the Working Group recommended that a preliminary TAC for the 2003/04 capelin fishery be set at 555000 t .

According to the results of the last of the surveys carried out in January-February 2004 (survey results described in section 5.3.2; Tables 5.3.2.1.2-5.3.2.1.4), the fishable spawning stock was estimated at $50.7 * 10^{9}$ fish on 15 February 2004. At that time the observed mean weight in the fishable stock was 18.3 g and the stock biomass was about 925000 t (Table 5.3.2.1.4). With the usual prerequisites of a monthly natural mortality rate of 0.035 and a remaining spawning stock of 400000 t , the above abundance estimate indicated that a TAC of 525000 t was available for the remainder of the 2004 winter fishery. Adding this to the catch already taken from June 2003 - mid-February 2004 gave a total TAC of $875000 t$ for the $2003 / 2004$ season. The TAC deviates from that predicted for the $2003 / 04$ season by 40000 t . This is explained by a $30 \%$ larger contribution in numbers of age group 2(3) (year class 2001) on one hand and somewhat lower mean weights than predicted.

About 135000 t of the TAC remained at the end of the winter fishery. As a result 535000 t of capelin remained to spawn in 2004 (Table 5.1.1).

### 5.5.2 Stock prognosis and assessment for the 2004/2005 season

As described in section 5.3.2.2., all attempts to locate and assess immature age 1 and 2 capelin (age 2 and 3 in 2004) were unsuccessful. In fact, only trace amounts were measured where surveys located any of these capelin at all. Therefore, there are no data available to the Working Group for predicting likely fishable/spawning stock size in the 2004/2005 season.

Under these circumstances, the Working Group recommends that a fishery is not opened, unless further surveys in 2004 and/or 2005 reveal large enough numbers and biomass of capelin of the 2001 and 2002 year classes to sustain a fishery and to meet the prerequisite of a remaining spawning stock of 400000 t in spring 2005.

### 5.5.3 Management of capelin in the Iceland-East Greenland-Jan Mayen area

The fishable stock consists of 2 age groups ( 2 and 3 year olds, spawning at ages 3 and 4 ). The fishing season usually begins in June/July and ends in March of the following year when the remainder of the fishable stock spawns and dies. The fishable stock, which is also the maturing stock, is thus renewed annually and its exploitation must of necessity be cautious. Due to the short life span and high spawning mortality, stock abundance can only be assessed by acoustic surveys.

Since 1992, the key elements in the management of capelin in the Iceland-East Greenland-Jan Mayen area have been as follows:

Acoustic survey estimates of juvenile capelin abundance have been used to predict fishable stock abundance in the following year (fishing season). Historical average mean weight at age (in later years a relationship between numerical stock abundance and growth), growth rates and natural mortality have been used for calculations and projections of maturing and fishable stock biomass.

Based on the data described above, a TAC is predicted in the spring of the year in which the season begins, allowing for $400000 t$ to spawn at the end of the season. For precautionary purposes, a preliminary TAC, corresponding to $2 / 3$ of the predicted total TAC for the season, has then been allocated to the period July-December. With regard to a precautionary approach, the Working Group stresses the importance of the continued setting of a preliminary TAC for the first half of the season.

The final decision on a TAC for each fishing season has been based on the results of acoustic stock abundance surveys in late autumn or in January/February of the following year during the fishing season.

The procedure just described has worked well in the past for 'normal' ranges of stock abundance. However, it is clear that extra care should be taken when dealing with stock abundance below or above the norm, corresponding to TACs lower than 500000 t or greater than 1600000 t .

### 5.6 Precautionary Approach to Fisheries Management

Due to the short life span of capelin and their high spawning mortality, the main management objective is to maintain enough spawners for the propagation of the stock. Since 1979 the targeted remaining spawning stock for capelin in the Iceland-East Greenland Jan Mayen area has been 400000 t . Although there have been large fluctuations in stock abundance during this period, these appear to be environmentally induced and not due to excessive fishing. Therefore, the criterion of maintaining a remaining spawning stock may be defined as $\mathbf{B}_{\mathrm{lim}}$, i.e. stock abundance below which no fishery should be permitted.

The definition of other precautionary reference points is more problematic. However, due to uncertainties inherent in predicting the abundance of short-lived species and the importance of capelin as forage fish for predators such as cod, saithe, Greenland halibut, baleen whales and sea birds, extra precaution should be taken when stock biomass projections indicate TACs lower than 500000 t and greater than 1600000 t . In the former case, the fishery should not be opened until after the completion of a stock assessment survey in autumn/winter in that season. The latter simply represents a scenario where projected stock abundance is beyond the highest historical abundance on record. In such cases the preliminary TAC should not exceed 1100000 t .

### 5.7 Special Comments

In the last two years great difficulties have been encountered in locating and assessing the juvenile part of the stock (ages 1 and 2; 2 and 3 after 31 December). In this period, the quarterly monitoring of environmental conditions of Icelandic waters, show a rise in sea temperatures north and east of Iceland, which probably also reaches farther north and northwest. The temperature increase is so large that it has probably led to displacements of the juvenile part of the capelin stock. On the basis of experience gained before and during assessment surveys of the 2003/2004 season, these displacements have obviously been so large that juvenile capelin, in particular the 2001 year class, were not available to the autumn 2003 assessment survey as it was carried out. Therefore, while the very low numbers of immature capelin of the 2001 and 2002 year classes hitherto recorded should be taken seriously, they do not necessarily indicate a radical decline of the adult fishable stock in the 2004/2005 season.

An overview of stock developments during 1978-2004 is given in Table 5.7.1.

| Investigation | No. of samples | Length meas. individuals | Aged individuals |
| :--- | :---: | :---: | :---: |
| Fishery 2003 | 15 | 1200 | 1465 |
| Survey 2003 | 15 | 1377 | 1320 |
| Fishery 2004 | 30 | 3000 | 2756 |
| Survey 2004 | 60 | 6000 | 5882 |

Table 5.1.1 Preliminary TACs for the summer/autumn fishery, recommended TACs for the entire season, landings and remaining spawning stock (000 tonnes) in the 1991/92-2003/04 seasons.

| Season | $91 / 92$ | $92 / 93$ | $93 / 94$ | $94 / 95$ | $95 / 96$ | $96 / 97$ | $97 / 98$ | $98 / 99$ | $99 / 00$ | $00 / 01$ | $01 / 02$ | $02 / 03$ | $03 / 04$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Prelim. TAC | 0 | 500 | 900 | 950 | 800 | 1100 | 850 | 950 | 866 | 975 | 1050 | 1040 | 835 |
| Rec. TAC | 740 | 900 | 1250 | 850 | 1390 | 1600 | 1265 | 1200 | 1000 | 1090 | 1325 | 1000 | 875 |
| Landings | 677 | 788 | 1179 | 842 | 930 | 1571 | 1245 | 1100 | 934 | 1065 | 1249 | 988 | 741 |
| Spawn. stock | 475 | 460 | 460 | 420 | 830 | 430 | 492 | 500 | 650 | 450 | 475 | 410 | 535 |

Table 5.2.1 The international capelin catch 1964-2004 (thousand tonnes).

| Year | Winter season |  |  |  |  | Summer and autumn season |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Iceland | Norway | Faroes | Greenland | $\begin{array}{r} \hline \text { Season } \\ \text { total } \end{array}$ | Iceland | Norway | Faroes | Greenland | EU | $\begin{array}{r} \hline \text { Season } \\ \text { total } \\ \hline \end{array}$ |  |
| 1964 | 8.6 | - | - |  | 8.6 | - | - | - |  | - | - | 8.6 |
| 1965 | 49.7 | - | - |  | 49.7 | - | - | - |  | - | - | 49.7 |
| 1966 | 124.5 | - | - |  | 124.5 | - | - | - |  | - | - | 124.5 |
| 1967 | 97.2 | - | - |  | 97.2 | - | - | - |  | - | - | 97.2 |
| 1968 | 78.1 | - | - |  | 78.1 | - | - | - |  | - | - | 78.1 |
| 1969 | 170.6 | - | - |  | 170.6 | - | - | - |  | - | - | 170.6 |
| 1970 | 190.8 | - | - |  | 190.8 | - | - | - |  | - | - | 190.8 |
| 1971 | 182.9 | - | - |  | 182.9 | - | - | - |  | - | - | 182.9 |
| 1972 | 276.5 | - | - |  | 276.5 |  | - | - |  | - | - | 276.5 |
| 1973 | 440.9 | - | - |  | 440.9 | - | - | - |  | - | - | 440.9 |
| 1974 | 461.9 | - | - |  | 461.9 | - | - | - |  | - | - | 461.9 |
| 1975 | 457.1 | - | - |  | 457.1 | 3.1 | - | - |  | - | 3.1 | 460.2 |
| 1976 | 338.7 | - | - |  | 338.7 | 114.4 | - | - |  | - | 114.4 | 453.1 |
| 1977 | 549.2 | - | 24.3 |  | 573.5 | 259.7 | - | - |  | - | 259.7 | 833.2 |
| 1978 | 468.4 | - | 36.2 |  | 504.6 | 497.5 | 154.1 | 3.4 |  | - | 655.0 | 1,159.6 |
| 1979 | 521.7 | - | 18.2 |  | 539.9 | 442.0 | 124.0 | 22.0 |  | - | 588.0 | 1,127.9 |
| 1980 | 392.1 | - | - |  | 392.1 | 367.4 | 118.7 | 24.2 |  | 17.3 | 527.6 | 919.7 |
| 1981 | 156.0 | - | - |  | 156.0 | 484.6 | 91.4 | 16.2 |  | 20.8 | 613.0 | 769.0 |
| 1982 | 13.2 | - | - |  | 13.2 | - | - | - |  | - | - | 13.2 |
| 1983 | - | - | - |  | - | 133.4 | - | - |  | - | 133.4 | 133.4 |
| 1984 | 439.6 | - | - |  | 439.6 | 425.2 | 104.6 | 10.2 |  | 8.5 | 548.5 | 988.1 |
| 1985 | 348.5 | - | - |  | 348.5 | 644.8 | 193.0 | 65.9 |  | 16.0 | 919.7 | 1,268.2 |
| 1986 | 341.8 | 50.0 | - |  | 391.8 | 552.5 | 149.7 | 65.4 |  | 5.3 | 772.9 | 1,164.7 |
| 1987 | 500.6 | 59.9 | - |  | 560.5 | 311.3 | 82.1 | 65.2 |  | - | 458.6 | 1,019.1 |
| 1988 | 600.6 | 56.6 | - |  | 657.2 | 311.4 | 11.5 | 48.5 |  | - | 371.4 | 1,028.6 |
| 1989 | 609.1 | 56.0 | - |  | 665.1 | 53.9 | 52.7 | 14.4 |  | - | 121.0 | 786,1 |
| 1990 | 612.0 | 62.5 | 12.3 |  | 686,8 | 83.7 | 21.9 | 5.6 |  | - | 111.2 | 798.0 |
| 1991 | 202.4 | - | - |  | 202.4 | 56.0 | - | - |  | - | 56.0 | 258.4 |
| 1992 | 573.5 | 47.6 | - |  | 621.1 | 213.4 | 65.3 | 18.9 | 0.5 | - | 298.1 | 919.2 |
| 1993 | 489.1 | - | - | 0.5 | 489.6 | 450.0 | 127.5 | 23.9 | 10.2 | - | 611.6 | 1,101.2 |
| 1994 | 550.3 | 15.0 | - | 1.8 | 567.1 | 210.7 | 99.0 | 12.3 | 2.1 | - | 324.1 | 891.2 |
| 1995 | 539.4 | - | - | 0.4 | 539.8 | 175.5 | 28.0 | - | 2.2 | - | 205.7 | 745.5 |
| 1996 | 707.9 | - | 10.0 | 5.7 | 723.6 | 474.3 | 206.0 | 17.6 | 15.0 | 60.9 | 773.8 | 1,497.4 |
| 1997 | 774.9 | - | 16.1 | 6.1 | 797.1 | 536.0 | 153.6 | 20.5 | 6.5 | 47.1 | 763.6 | 1,561.5 |
| 1998 | 457.0 | - | 14.7 | 9.6 | 481.3 | 290.8 | 72.9 | 26.9 | 8.0 | 41.9 | 440.5 | 921.8 |
| 1999 | 607.8 | 14.8 | 13.8 | 22.5 | 658.9 | 83.0 | 11.4 | 6.0 | 2.0 | - | 102.4 | 761.3 |
| 2000 | 761.4 | 14.9 | 32.0 | 22.0 | 830.3 | 126.5 | 80.1 | 30.0 | 7.5 | 21.0 | 265.1 | 1,095.4 |
| 2001 | 767.2 | - | 10.0 | 29.0 | 806.2 | 150.0 | 106.0 | 12.0 | 9.0 | 17.0 | 294.0 | 1,061.2 |
| 2002 | 901.0 | - | 28.0 | 26.0 | 955.0 | 180.0 | 118.7 | - | 13.0 | 28.0 | 339.7 | 1,294.7 |
| 2003 | 585.0 | - | 40.0 | 23.0 | 648.0 | 96.5 | 78.0 | 3.5 | 2.5 | 18.0 | 198.5 | 846.5 |
| 2004 | 478.8 | 15.8 | 30.8 | 17.5 | 542.9 |  |  |  |  |  |  |  |

Table 5.2.2 The total international catch of capelin in the Iceland-East Greenland-Jan Mayen area by age group in numbers (billions) and the total catch by numbers and weight (thousand tonnes) in the autumn season (August-December) 1982-2003.

|  |  |  | Year |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 |
| 1 | - | 0.6 | 0.5 | 0.8 | + | + | 0.3 | 1.7 | 0.8 | 0.3 | 1.7 |
| 2 | - | 7.2 | 9.8 | 25.6 | 10.0 | 27.7 | 13.6 | 6.0 | 5.9 | 2.7 | 14.0 |
| 3 | - | 0.8 | 7.8 | 15.4 | 23.3 | 6.7 | 5.4 | 1.5 | 1.0 | 0.4 | 2.1 |
| 4 | - | - | 0.1 | 0.2 | 0.5 | + | + | + | + | + | + |
| Total number | - | 8.6 | 18.2 | 42.0 | 33.8 | 34.4 | 19.3 | 9.2 | 7.7 | 3.4 | 17.8 |
| Total weight | - | 133.4 | 548.5 | 919.7 | 772.9 | 458.6 | 371.4 | 121.0 | 111.2 | 56.0 | 298.1 |


|  |  |  | Year |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| 1 | 0.2 | 0.6 | 1.5 | 0.2 | 1.8 | 0.9 | 0.3 | 0.2 | + | + | 0.3 |
| 2 | 24.9 | 15.0 | 9.7 | 25.2 | 33.4 | 25.1 | 4.7 | 12.9 | 17.6 | 18.3 | 12.9 |
| 3 | 5.4 | 2.8 | 1.1 | 12.7 | 10.2 | 2.9 | 0.7 | 3.3 | 1.2 | 2.5 | 1.1 |
| 4 | 0.2 | + | + | 0.2 | 0.4 | + | + | 0.1 | + | + | + |
| Total number | 30.7 | 18.4 | 12.3 | 38.4 | 45.8 | 28.9 | 5.7 | 16.5 | 18.8 | 20.8 | 14.3 |
| Total weight | 611.6 | 324.1 | 205.7 | 773.7 | 763.6 | 440.5 | 102.4 | 265.1 | 294.0 | 339.7 | 199.5 |

Table 5.2.3 The total international catch of capelin in the Iceland-East Greenland-Jan Mayen area by age group in numbers (billions) and the total catch by numbers and weight (thousand tonnes) in the winter season (January-March) 1983-2004.

|  |  |  | Year |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| 2 | - | 2.1 | 0.4 | 0.1 | + | + | 0.1 | 1.4 | 0.5 | 2.7 | 0.2 |
| 3 | - | 18.1 | 9.1 | 9.8 | 6.9 | 23.4 | 22.9 | 24.8 | 7.4 | 29.4 | 20.1 |
| 4 | - | 3.4 | 5.4 | 6.9 | 15.5 | 7.2 | 7.8 | 9.6 | 1.5 | 2.8 | 2.5 |
| 5 | - | - | - | 0.2 | - | 0.3 | + | 0.1 | + | + | + |
| Total number | - | 23.6 | 14.5 | 17.0 | 22.4 | 30.9 | 30.8 | 35.9 | 9.4 | 34.9 | 22.8 |
| Total weight | - | 439.6 | 348.5 | 391.8 | 560.5 | 657.2 | 665.1 | 686.8 | 202.4 | 621.1 | 489.6 |


|  |  | Year |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 2 | 0.6 | 1.3 | 0.6 | 0.9 | 0.3 | 0.5 | 0.3 | 0.4 | 0.1 | 0.1 | 0.6 |
| 3 | 22.7 | 17.6 | 27.4 | 29.1 | 20.4 | 31.2 | 36.3 | 27.9 | 33.1 | 32.2 | 24.6 |
| 4 | 3.9 | 5.9 | 7.7 | 11.0 | 5.4 | 7.5 | 5.4 | 6.7 | 4.2 | 1.9 | 3.0 |
| 5 | + | + | + | + | + | + | + | + | + | + | + |
| Total number | 27.2 | 24.8 | 35.7 | 41.0 | 26.1 | 39.2 | 42.0 | 35.0 | 37.4 | 34.4 | 28.2 |
| Total weight | 567.1 | 539.8 | 723.6 | 797.6 | 481.3 | 658.9 | 830.3 | 806.2 | 955.0 | 648.0 | 542.9 |

Table 5.2.4 The total international catch in numbers (millions) of capelin in the Iceland-East GreenlandJan Mayen area in the summer/autumn season of 2003 by age and length, and the catch in weight (thousand tonnes) by age group.

| Total length (cm) | Age 1 | Age 2 | Age 3 | Age 4 | Total | Percentage |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |
|  | 9.5 | 57 | 0 | 0 |  | 57 |
| 10 | 0 | 0 | 0 |  | 0 | 0.4 |
|  | 11 | 86 | 0 | 0 | 0 | 86 |
| 11.5 | 0 | 29 | 0 | 0 | 29 | 0.6 |
| 12 | 57 | 258 | 0 | 0 | 315 | 2.2 |
| 12.5 | 57 | 315 | 0 | 0 | 372 | 2.6 |
| 13 | 29 | 1432 | 0 | 0 | 1461 | 10.2 |
|  | 0 | 2320 | 0 | 0 | 2320 | 16.2 |
|  | 13.5 | 0 | 3322 | 29 | 0 | 3351 |
|  | 0 | 2578 | 172 | 0 | 2749 | 19.4 |
|  | 14.5 | 0 | 1518 | 315 | 0 | 1833 |
| 15 | 0 | 859 | 229 | 0 | 1088 | 12.8 |
|  | 15.5 | 0 | 200 | 258 | 0 | 458 |
| Total number | 16 | 0 | 86 | 86 | 0 | 172 |
| Percentage | 0 | 0 | 29 | + | 29 | 1.2 |
| Total weight |  |  |  |  |  |  |

Table 5.2.5 The total international catch in numbers (millions) of capelin in the Iceland-East Greenland-Jan Mayen area in the winter season of 2004 by age and length, and the catch in weight (thousand tonnes) by age group.

| Total length (cm) | Age 2 | Age 3 | Age 4 | Age 5 | Total | Percentage |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 11 | 67 | 0 | 0 | 0 | 67 | 0.2 |
| 11.5 | 55 | 0 | 0 | 0 | 55 | 0.2 |
| 12 | 86 | 34 | 0 | 0 | 120 | 0.4 |
| 12.5 | 86 | 302 | 0 | 0 | 387 | 1.4 |
| 13 | 89 | 1564 | 0 | 0 | 1652 | 5.9 |
| 13.5 | 102 | 2398 | 0 | 0 | 2499 | 8.9 |
| 14 | 45 | 3037 | 23 | 0 | 3104 | 11.0 |
| 14.5 | 11 | 4143 | 197 | 0 | 4351 | 15.5 |
| 15 | 11 | 3779 | 400 | 0 | 4190 | 14.9 |
| 15.5 | 11 | 3669 | 461 | 0 | 4141 | 14.7 |
| 16 | 11 | 2875 | 703 | 11 | 3601 | 12.8 |
| 16.5 | 0 | 1606 | 539 | 0 | 2145 | 7.6 |
| 17 | 0 | 792 | 345 | 0 | 1137 | 4.0 |
| 17.5 | 0 | 320 | 258 | 0 | 578 | 2.1 |
| 18 | 0 | 76 | 28 | 0 | 104 | 0.4 |
| 18.5 | 0 | 0 | 12 | 0 | 12 | 0.0 |
| 19 | 0 | 0 | 0 | 0 | 0 | 0.0 |
| 19.5 | 0 | 0 | 12 | 0 | 12 | 0.0 |
| Total number | 574 | 24594 | 2978 | 11 | 28157 |  |
| Percentage | 2.0 | 87.3 | 10.6 | 0.0 | 100.0 | 100.0 |
| Total weight | 11.2 | 474.1 | 57.5 | 0.2 | 542.9 |  |

Table 5.3.1.1 Abundance indices of 0-group capelin 1970-2003 and their division by areas.

|  |  |  |  | Year |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Area | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 |
| NW-Irminger Sea | 1 | + | + | 14 | 26 | 3 | 2 | 2 | + | 4 | 3 | 10 | + |
| W-Iceland | 8 | 7 | 30 | 39 | 44 | 37 | 5 | 19 | 2 | 19 | 18 | 13 | 8 |
| N-Iceland | 2 | 12 | 52 | 46 | 57 | 46 | 10 | 19 | 29 | 25 | 19 | 6 | 5 |
| East Iceland | - | + | 7 | 17 | 7 | 3 | 15 | 3 | + | 1 | + | - | + |
| Total | 11 | 19 | 89 | 116 | 134 | 89 | 32 | 43 | 31 | 49 | 40 | 29 | 13 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  | Year |  |  |  |  |  |  |
| Area | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| NW-Irminger Sea | + | + | 1 | + | 1 | 3 | 1 | + | 8 | 3 | 2 | 3 | + |
| W-Iceland | 3 | 2 | 8 | 16 | 6 | 22 | 13 | 7 | 2 | 11 | 21 | 12 | 6 |
| N-Iceland | 18 | 17 | 19 | 17 | 6 | 26 | 24 | 12 | 43 | 20 | 13 | 69 | 10 |
| East Iceland | 1 | 9 | 3 | 4 | 1 | 1 | 2 | 2 | 1 | + | 15 | 10 | 8 |
| Total | 22 | 28 | 31 | 37 | 14 | 52 | 40 | 21 | 54 | 34 | 51 | 94 | 24 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  | Year |  |  |  |  |  |  |
|  | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |  |  |  |  |  |
| NW-Irminger Sea | 2 | 5 | + | NA | NA | NA | NA | NA |  |  |  |  |  |
| W-Iceland | 17 | 14 | 7 | 25 | 1 | 25 | + | 17 |  |  |  |  |  |
| N-Iceland | 57 | 30 | 34 | 51 | 7 | 53 | 4 | 8 |  |  |  |  |  |
| East Iceland | 6 | 12 | 5 | 7 | 4 | 4 | + | 1 |  |  |  |  |  |
| Total | 82 | 61 | 46 | 83 | 12 | 82 | 5 | 26 |  |  |  |  |  |

Table 5.3.1.2 Estimated numbers, mean length and weight of age 1 capelin in the August surveys for 1983-2001.

|  |  |  |  |  |  |  | Year |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| Number ( $10^{9}$ ) | 155 | 286 | 31 | 71 | 101 | 147 | 111 | 36 | 50 | 87 | 33 | 85 | 189 | 138 |
| Mean length (cm) | 10.4 | 9.7 | 10.2 | 9.5 | 9.1 | 8.8 | 10.1 | 10.4 | 10.7 | 9.7 | 9.4 | 9.0 | 9.8 | 9.3 |
| Mean weight (g) | 4.2 | 3.6 | 3.8 | 3.3 | 3.0 | 2.6 | 3.4 | 4.0 | 5.1 | 3.4 | 3.0 | 2.8 | 3.4 | 2.9 |
|  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | Year |  |  |  |  |  |  |  |
| Number ( $10^{9}$ ) | 143 | 87 | 55 | 94 | 99 | No survey |  |  |  |  |  |  |  |  |
| Mean length (cm) | 9.3 | 9.0 | 9.5 | 9.5 | 10.0 |  |  |  |  |  |  |  |  |  |
| Mean weight (g) | 2.8 | 2.9 | 3.2 | 3.1 | 3.7 |  |  |  |  |  |  |  |  |  |

Table 5.3.2.1.1 Acoustic abundance estimate of immature capelin, 6-21 November 2003

| Length | NUMBERS $\left(10^{-6}\right)$ |  |  |  | Avg wt <br> (g) | BIOMASS $\left(10^{-9}\right)$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | (Age) Year class |  |  | Total |  | (Age) Year class |  |  | Total |
|  | (1) 2002 | (2) 2001 | (3) 2000 |  |  | (1) 2002 | (2) 2001 | (3) 2000 |  |
| 9 | 200 | 0 | 0 | 200 | 3.1 | 0.6 | 0.0 | 0.0 | 0.6 |
| 9.5 | 200 | 0 | 0 | 200 | 3.6 | 0.7 | 0.0 | 0.0 | 0.7 |
| 10 | 200 | 0 | 0 | 200 | 4.2 | 0.8 | 0.0 | 0.0 | 0.8 |
| 10.5 | 200 | 0 | 0 | 200 | 9.0 | 1.8 | 0.0 | 0.0 | 1.8 |
| 11 | 600 | 0 | 0 | 600 | 5.3 | 3.2 | 0.0 | 0.0 | 3.2 |
| 11.5 | 1200 | 0 | 0 | 1200 | 6.2 | 7.4 | 0.0 | 0.0 | 7.4 |
| 12 | 700 | 100 | 0 | 800 | 7.2 | 5.0 | 0.7 | 0.0 | 5.8 |
| 12.5 | 700 | 100 | 0 | 800 | 8.2 | 5.7 | 0.8 | 0.0 | 6.6 |
| 13 | 300 | 200 | 0 | 600 | 9.3 | 2.8 | 1.9 | 0.0 | 4.7 |
| 13.5 | 200 | 300 | 0 | 400 | 10.7 | 2.1 | 3.2 | 0.0 | 5.4 |
| 14 | 0 | 400 | 0 | 500 | 12.5 | 0.0 | 5.0 | 0.0 | 5.0 |
| 14.5 | 100 | 400 | 0 | 500 | 13.9 | 1.4 | 5.6 | 0.0 | 7.0 |
| 15 | 0 | 600 | 0 | 600 | 16.0 | 0.0 | 9.6 | 0.0 | 9.6 |
| 15.5 | 0 | 400 | 0 | 500 | 17.8 | 0.0 | 7.1 | 0.0 | 7.1 |
| 16 | 0 | 200 | 0 | 300 | 20.5 | 0.0 | 4.1 | 0.0 | 4.1 |
| 16.5 | 0 | 200 | 100 | 200 | 23.0 | 0.0 | 4.6 | 2.3 | 6.9 |
| 17 | 0 | 0 | 0 | 0 | 25.4 | 0.0 | 0.0 | 0.0 | 0.0 |
| 17.5 | 0 | 0 | 0 | 0 | 27.6 | 0.0 | 0.0 | 0.0 | 0.0 |
| 18 | 0 | 0 | 0 | 0 | 32.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Total | 4600 | 3100 | 100 | 7800 |  | 31.7 | 42.6 | 2.3 | 76.6 |
| Average length |  |  |  |  |  | 11.3 | 15.7 | 16.7 | 15.7 |
| Average weight |  |  |  |  |  | 6.9 | 17.7 | 22.3 | 17.9 |

Table 5.3.2.1.2 Acoustic abundance estimate of mature capelin, 3-12 January, 2004

| $\begin{aligned} & \text { Length } \\ & (\mathrm{cm}) \\ & \hline \end{aligned}$ | NUMBERS ( $10^{-6}$ ) |  |  |  | Avg wt <br> (g) | BIOMASS ( $10^{-9}$ ) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | (Age) Year class |  |  | Total |  | (Age) Year class |  |  | Total |
|  | (2) 2002 | (3) 2001 | (4) 2000 |  |  | (2) 2002 | (3) 2001 | (4) 2000 |  |
| 12 | 1 | 0 | 0 | 1 | 6.2 | 0.0 | 0.0 | 0.0 | 0.0 |
| 12.5 | 10 | 0 | 0 | 10 | 7.4 | 0.1 | 0.0 | 0.0 | 0.1 |
| 13 | 10 | 0 | 0 | 10 | 8.3 | 0.1 | 0.0 | 0.0 | 0.1 |
| 13.5 | 10 | 200 | 0 | 210 | 9.6 | 0.1 | 1.9 | 0.0 | 2.0 |
| 14 | 10 | 500 | 0 | 510 | 11.0 | 0.1 | 5.5 | 0.0 | 5.6 |
| 14.5 | 0 | 800 | 0 | 800 | 12.5 | 0.0 | 10.0 | 0.0 | 10.0 |
| 15 | 0 | 1200 | 0 | 1200 | 14.8 | 0.0 | 17.8 | 0.0 | 17.8 |
| 15.5 | 0 | 1200 | 0 | 1200 | 16.6 | 0.0 | 19.9 | 0.0 | 19.9 |
| 16 | 0 | 1000 | 100 | 1100 | 19.0 | 0.0 | 19.0 | 1.9 | 20.9 |
| 16.5 | 0 | 800 | 105 | 905 | 21.8 | 0.0 | 17.4 | 2.2 | 19.6 |
| 17 | 0 | 600 | 100 | 700 | 24.5 | 0.0 | 14.7 | 2.5 | 17.2 |
| 17.5 | 0 | 300 | 50 | 350 | 27.4 | 0.0 | 8.2 | 1.4 | 9.6 |
| 18 | 0 | 100 | 10 | 110 | 29.1 | 0.0 | 2.9 | 0.3 | 3.2 |
| 18.5 | 0 | 0 | 30 | 30 | 32.7 | 0.0 | 0.0 | 1.0 | 1.0 |
| 19 | 0 | 0 | 5 | 5 | 33.8 | 0.0 | 0.0 | 0.2 | 0.2 |
| Total | 41 | 6700 | 400 | 7141 |  | 0.4 | 117.4 | 9.3 | 127.1 |
| Average length |  |  |  |  |  | 13.2 | 15.6 | 16.8 | 15.6 |
| Average weight |  |  |  |  |  | 9.4 | 17.6 | 23.6 | 17.8 |

Table 5.3.2.1.3 Acoustic abundance estimate of mature capelin, 18-30 January, 2004

| Length (cm) | NUMBERS ( $10^{-6}$ ) |  |  |  | Avg wt <br> (g) | BIOMASS ( $10^{-9}$ ) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | (Age) Year class |  |  | Total |  | (Age) Year class |  |  | Total |
|  | (2) 2002 | (3) 2001 | (4) 2000 |  |  | (2) 2002 | (3) 2001 | (4) 2000 |  |
| 12.5 | 1 | 0 | 0 | 1 | 6.9 | 0.1 | 0.0 | 0.0 | 0.1 |
| 13 | 50 | 200 | 0 | 200 | 8.3 | 0.2 | 1.7 | 0.0 | 1.9 |
| 13.5 | 200 | 700 | 0 | 900 | 10.0 | 2.0 | 7.0 | 0.0 | 9.0 |
| 14 | 100 | 2200 | 100 | 2400 | 11.1 | 1.1 | 24.4 | 1.1 | 26.6 |
| 14.5 | 30 | 4100 | 0 | 4200 | 12.7 | 0.4 | 52.1 | 0.0 | 52.5 |
| 15 | 0 | 4400 | 100 | 4500 | 14.3 | 0.0 | 62.9 | 1.4 | 64.4 |
| 15.5 | 0 | 5400 | 200 | 5600 | 16.2 | 0.0 | 87.5 | 3.2 | 90.7 |
| 16 | 0 | 5700 | 200 | 5900 | 18.6 | 0.0 | 106.0 | 3.7 | 109.7 |
| 16.5 | 0 | 4700 | 300 | 5000 | 20.9 | 0.0 | 98.2 | 6.3 | 104.5 |
| 17 | 0 | 4900 | 200 | 5200 | 23.6 | 0.0 | 115.6 | 4.7 | 120.4 |
| 17.5 | 0 | 3700 | 700 | 4400 | 26.5 | 0.0 | 98.1 | 18.6 | 116.6 |
| 18 | 0 | 1800 | 400 | 2100 | 29.2 | 0.0 | 52.6 | 11.7 | 64.2 |
| 18.5 | 0 | 500 | 100 | 600 | 31.6 | 0.0 | 15.8 | 3.2 | 19.0 |
| 19 | 0 | 0 | 5 | 5 | 30.6 | 0.0 | 0.0 | 0.2 | 0.2 |
| Total | 381 | 38300 | 2305 | 41006 |  | 3.8 | 721.9 | 54.0 | 779.7 |
| Average length |  |  |  |  |  | 13.8 | 15.9 | 16.9 | 16.0 |
| Average weight |  |  |  |  |  | 10.7 | 18.9 | 23.5 | 19.0 |

Table 5.3.2.1.4 Acoustic abundance estimate of mature capelin, 10-18 February, 2004

| $\begin{aligned} & \text { Length } \\ & (\mathrm{cm}) \end{aligned}$ | NUMBERS ( $10^{-6}$ ) |  |  |  | Avg wt <br> (g) | BIOMASS ( $10^{-9}$ ) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | (Age) Year class |  |  | Total |  | (Age) Year class |  |  | Total |
|  | (2) 2002 | (3) 2001 | (4) 2000 |  |  | (2) 2002 | (3) 2001 | (4) 2000 |  |
| 12 | 367 | 0 | 0 | 367 | 6.9 | 2.5 | 0.0 | 0.0 | 2.5 |
| 12.5 | 226 | 84 | 0 | 310 | 8.8 | 2.0 | 0.8 | 0.0 | 2.7 |
| 13 | 536 | 591 | 0 | 1126 | 9.9 | 5.3 | 5.8 | 0.0 | 11.1 |
| 13.5 | 734 | 2280 | 0 | 3015 | 11.1 | 8.1 | 25.3 | 0.0 | 33.5 |
| 14 | 1070 | 3803 | 28 | 4902 | 12.8 | 13.7 | 48.8 | 0.4 | 62.9 |
| 14.5 | 592 | 5435 | 84 | 6110 | 14.5 | 8.7 | 79.0 | 1.2 | 88.8 |
| 15 | 311 | 7605 | 480 | 8394 | 16.7 | 5.2 | 127.0 | 8.0 | 140.2 |
| 15.5 | 170 | 7717 | 816 | 8701 | 18.7 | 3.2 | 144.5 | 15.3 | 162.9 |
| 16 | 84 | 6590 | 872 | 7548 | 21.1 | 1.8 | 139.4 | 18.4 | 159.6 |
| 16.5 | 0 | 4226 | 620 | 4847 | 24.1 | 0.0 | 101.8 | 15.0 | 116.7 |
| 17 | 0 | 2534 | 986 | 3520 | 26.2 | 0.0 | 66.2 | 26.0 | 92.2 |
| 17.5 | 0 | 704 | 479 | 1183 | 29.2 | 0.0 | 20.3 | 14.1 | 34.5 |
| 18 | 0 | 84 | 282 | 367 | 33.3 | 0.0 | 2.8 | 9.4 | 12.2 |
| 18.5 | 0 | 84 | 84 | 198 | 34.1 | 0.0 | 2.8 | 2.8 | 6.8 |
| Total | 4198 | 41739 | 4732 | 50698 |  | 50.6 | 764.5 | 110.6 | 926.8 |
| Average length |  |  |  |  |  | 14.2 | 15.3 | 16.3 | 15.2 |
| Average weight |  |  |  |  |  | 13.8 | 18.3 | 23.3 | 18.3 |

Table 5.4.1 The estimated number (billions) of capelin on 1 August 1978-2003 by age and maturity groups. The total number (billions) and weight (thousand tonnes) of the immature and maturing (fishable) stock components are also given.

|  | Year |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age/maturity | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 |
| 1 juvenile | 163.8 | 60.3 | 66.1 | 48.9 | 146.4 | 124.2 | 250.5 | 98.9 | 156.2 | 144.0 |
| 2 immature | 15.3 | 16.4 | 4.2 | 3.7 | 15.0 | 42.5 | 40.9 | 100.0 | 29.4 | 37.2 |
| 2 mature | 81.9 | 91.3 | 35.4 | 39.7 | 17.1 | 53.7 | 40.7 | 64.6 | 35.6 | 65.4 |
| 3 mature | 29.1 | 10.1 | 10.8 | 2.8 | 2.3 | 9.8 | 27.9 | 27.0 | 65.8 | 20.1 |
| 4 mature | 0.4 | 0.3 | + | + | + | 0.1 | 0.4 | 0.4 | 0.7 | 0.1 |
| Number immat. | 179.2 | 76.7 | 70.3 | 52.6 | 161.4 | 166.7 | 291.4 | 198.9 | 185.6 | 181.2 |
| Number mature | 111.4 | 101.7 | 46.2 | 42.5 | 19.4 | 63.6 | 69.0 | 92.0 | 102.1 | 85.6 |
| Weight immat | 751 | 366 | 283 | 209 | 683 | 985 | 1067 | 1168 | 876 | 950 |
| Weight mature | 2081 | 1769 | 847 | 829 | 355 | 1085 | 1340 | 1643 | 2260 | 1689 |


|  |  |  |  | Year |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age/maturity | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| 1 juvenile | 80.8 | 63.9 | 117.5 | 132.9 | 162.9 | 144.3 | 224.1 | 197.3 | 191.2 | 165.4 |
| 2 immature | 24.0 | 10.3 | 10.1 | 9.7 | 16.6 | 20.1 | 35.2 | 45.1 | 28.7 | 35.2 |
| 2 mature | 70.3 | 42.8 | 31.9 | 67.7 | 70.7 | 86.9 | 59.8 | 102.2 | 100.7 | 90.3 |
| 3 mature | 24.5 | 15.8 | 6.8 | 6.7 | 6.4 | 10.9 | 13.2 | 23.0 | 29.6 | 19.0 |
| 4 mature | 0.4 | + | + | + | + | 0.2 | - | + | + | + |
| Number immat. | 104.8 | 74.2 | 127.6 | 142.6 | 179.5 | 164.7 | 259.2 | 242.4 | 219.9 | 200.6 |
| Number mature | 95.2 | 58.6 | 38.7 | 74.4 | 77.1 | 98.0 | 73.0 | 125.1 | 130.3 | 109.3 |
| Weight immat | 438 | 309 | 542 | 702 | 747 | 702 | 1019 | 1188 | 985 | 758 |
| Weight mature | 1663 | 1173 | 751 | 1273 | 1311 | 1585 | 1268 | 1819 | 1900 | 1590 |


|  |  |  | Year |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age/maturity | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 1 juvenile | 167.9 | 138.0 | 145.6 | 139.7 | $165.9^{*}$ | NA |  |
| 2 immature | 19.2 | 24.4 | 25.0 | 9.0 | 23.9 | $26.9^{* *}$ |  |
| 2 mature | 89.5 | 85.9 | 65.7 | 86.7 | 68.0 | 82.1 | NA |
| 3 mature | 23.2 | 12.6 | 16.0 | 16.9 | 5.9 | 15.7 | $17.7^{* *}$ |
| 4 mature | + | + |  |  |  |  |  |
| Number immat. | 187.1 | 162.4 | 170.6 | 148.7 | $189.8^{*}$ | $26.9^{* * *}$ |  |
| Number mature | 112.7 | 98.5 | 81.7 | 103.6 | 73.9 | 97.8 | $17.7^{* * *}$ |
| Weight immat | 621 | 612 | 645 | 615 | $877^{*}$ | $415^{* * *}$ |  |
| Weight mature | 1576 | 1703 | 1519 | 1817 | 1280 | 1544 | $410^{* * *}$ |

[^1]NA: Not available

Table 5.4.2 The estimated number (billions) of capelin on 1 January 1979-2004 by age and maturity groups. The total number (billions) and weight (thousand tonnes) of the immature and maturing (fishable) stock components and the remaining spawning stock by number and weight are also given.

|  | Year |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age/maturity | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| 2 juvenile | 137.6 | 50.6 | 55.3 | 41.2 | 123.7 | 105.0 | 211.6 | 83.2 | 131.9 | 120.5 |
| 3 immature | 12.8 | 13.8 | 3.5 | 3.0 | 12.6 | 35.7 | 34.3 | 83.9 | 25.6 | 31.2 |
| 3 mature | 51.8 | 53.4 | 16.3 | 8.0 | 14.3 | 39.8 | 25.2 | 34.5 | 22.1 | 34.1 |
| 4 mature | 14.8 | 3.6 | 4.9 | 0.5 | 2.0 | 7.6 | 15.6 | 10.5 | 37.0 | 11.7 |
| 5 mature | 0.3 | 0.2 | + | + | + | 0.1 | 0.3 | 0.2 | 0.2 | + |
| Number immat. | 150.4 | 64.4 | 58.8 | 44.2 | 136.3 | 140.7 | 245.9 | 167.1 | 157.5 | 151.3 |
| Number mature | 66.9 | 57.2 | 21.2 | 8.5 | 16.3 | 47.5 | 41.1 | 45.2 | 59.1 | 45.8 |
| Weight immat. | 1028 | 502 | 527 | 292 | 685 | 984 | 1467 | 1414 | 1003 | 1083 |
| Weight mature | 1358 | 980 | 471 | 171 | 315 | 966 | 913 | 1059 | 1355 | 993 |
| Number sp.st. | 29.0 | 17.5 | 7.7 | 6.8 | 13.5 | 21.6 | 20.7 | 19.6 | 18.3 | 18.5 |
| Weight sp. st | 600 | 300 | 170 | 140 | 260 | 440 | 460 | 460 | 420 | 400 |


|  |  |  | Year |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age/maturity | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| 2 juvenile | 67.8 | 53.9 | 98.9 | 111.6 | 124.6 | 121.3 | 188.1 | 165.2 | 160.0 | 138.8 |
| 3 immature | 20.1 | 8.6 | 8.6 | 8.1 | 13.9 | 16.9 | 29.5 | 37.9 | 24.1 | 29.5 |
| 3 mature | 48.8 | 31.2 | 22.3 | 54.8 | 46.5 | 50.5 | 35.1 | 75.5 | 72.4 | 50.1 |
| 4 mature | 16.0 | 12.1 | 4.5 | 5.3 | 3.5 | 4.6 | 8.7 | 20.1 | 24.8 | 7.9 |
| 5 mature | 0.3 | + | + | + | + | + | + | + | + | + |
| Number immat. | 87.9 | 62.5 | 107.5 | 119.7 | 138.5 | 138.2 | 217.6 | 203.1 | 184.1 | 168.3 |
| Number mature | 64.8 | 43.3 | 26.8 | 60.1 | 50.0 | 55.1 | 43.8 | 95.6 | 97.2 | 58.0 |
| Weight immat. | 434 | 291 | 501 | 487 | 622 | 573 | 925 | 800 | 672 | 621 |
| Weight mature | 1298 | 904 | 544 | 1106 | 1017 | 1063 | 914 | 1820 | 1881 | 1106 |
| Number sp.st. | 22.0 | 5.5 | 16.3 | 25.8 | 23.6 | 24.8 | 19.2 | 42.8 | 21.8 | 27.6 |
| Weight sp. st. | 440 | 115 | 330 | 475 | 499 | 460 | 420 | 830 | 430 | 492 |


|  |  | Year |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Age/maturity | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 2 juvenile | 140.9 | 115.8 | 122.2 | 117.3 | $139.2^{*}$ | NA |
| 3 immature | 16.1 | 20.5 | 21.0 | 7.6 | 20.1 | $22.6^{*}$ |
| 3 mature | 53.2 | 68.2 | 46.3 | 59.3 | 58.4 | 54.2 |
| 4 mature | 16.0 | 10.0 | 10.5 | 10.5 | 2.9 | 6.2 |
| 5 mature | + | + | + | + | + | + |
| Number immat. | 157.0 | 136.3 | 143.2 | 124.9 | $159.3^{*}$ | $22.6^{* *}$ |
| Number mature | 69.3 | 78.2 | 56.8 | 69.8 | 61.3 | 60.4 |
| Weight immat. | 585 | 535 | 655 | 510 | $750^{*}$ | $315^{* *}$ |
| Weight mature | 1171 | 1485 | 1197 | 1445 | 1214 | 1204 |
| Number sp.st. | 29.5 | 34.2 | 21.3 | 22.9 | 20.7 | 28.2 |
| Weight sp. st. | 500 | 650 | 450 | 475 | 410 | 535 |
|  |  |  |  |  |  |  |

*Preliminary/Predicted
** Year class 2001 only
NA: Not available

Table 5.5.1.1 Data used in the comparisons between abundance of age groups (numbers) when predicting fishable stock abundance for the calculation of preliminary TACs.

|  | Age 1 <br> Acoustics | Age 2 <br> Back-calc. <br> Mature | Age 2 <br> Acoustics <br> Immature | Age 2 <br> Back-calc. <br> Total | Age 3 <br> Back-calc. <br> Mature |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year |  | $\mathrm{N}_{1}$ | $\mathrm{~N}_{2 \text { mat }}$ | $\mathrm{N}_{2 \text { imm }}$ | $\mathrm{N}_{2 \text { tot }}$ |

* Invalid due to ice conditions.
** Preliminary
*** Projected from assessment in April 2003
NA: Not available

Table 5.5.1.2 Mean weight $(\mathrm{g})$ in autumn of maturing capelin.

|  |  |  | Years |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| Age 2 | 19.2 | 16.5 | 16.1 | 15.8 | 15.5 | 18.1 | 17.9 | 15.5 |
| Age 3 | 24.0 | 24.1 | 22.5 | 25.7 | 23.8 | 24.1 | 25.8 | 23.4 |
|  |  |  |  |  | Years |  |  |  |
|  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| Age 2 | 18.0 | 18.1 | 16.3 | 16.5 | 16.2 | 16.0 | 15.3 | 15.8 |
| Age 3 | 25.5 | 25.5 | 25.4 | 22.6 | 23.3 | 23.6 | 20.5 | 20.6 |
|  |  |  |  |  |  |  |  |  |
|  | 1997 | 1998 | 1999 | Years |  |  |  |  |
| Age 2 | 14.3 | 14.1 | 16.8 | 17.1 | 16.3 | No data | No data |  |
| Age 3 | 20.3 | 18.1 | 20.6 | 24.7 | 23.9 | No data | No data |  |

Table 5.5.1.3 Predictions of fishable stock abundance and TACs for the 1986/87-2003/04 seasons.
The last row gives contemporary advice on TACs for comparison.
Age 2 and age $3=$ Numbers in billions in age groups at the beginning of season.
Fishable stock = calculated weight of maturing capelin in thousand tonnes (ref. August).
TAC calc $=$ predicted in thousand tonnes.

| Season | $86 / 87$ | $87 / 88$ | $88 / 89$ | $89 / 90$ | $90 / 91$ | $91 / 92$ | $92 / 93$ | $93 / 94$ | $94 / 95$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year classes | $84-83$ | $85-84$ | $86-85$ | $87-86$ | $88-87$ | $89-88$ | $90-89$ | $91-90$ | $92-91$ |
| Age 2 | 34.9 | 55.5 | 64.8 | 43.2 | 31.1 | 39.4 | 56.4 | 93.1 | 89.6 |
| Age 3 | 55.0 | 13.7 | 29.0 | 25.5 | 8.2 | 3.7 | 18.3 | 22.6 | 27.0 |
| Fishable stock | 1926 | 1268 | 1800 | 1350 | 724 | 755 | 1398 | 2123 | 2170 |
| Calculated TAC | 1215 | 642 | 1105 | 713 | 170 | 197 | 755 | 1385 | 1427 |
| Advised TAC | 1333 | 1115 | 1036 | 550 | 265 | 740 | $* 900$ | 1250 | 850 |
|  |  |  |  |  |  |  |  |  |  |
| Season | $95 / 96$ | $96 / 97$ | $97 / 98$ | $98 / 99$ | $99 / 00$ | $00 / 01$ | $01 / 02$ | $02 / 03$ | $03 / 04$ |
| Year classes | $93-92$ | $94-93$ | $95-94$ | $96-95$ | $97 / 96$ | $98 / 97$ | $99 / 98$ | $00 / 99$ | $01 / 02$ |
| Age 2 | 92.5 | 90.0 | 83.8 | 94.4 | 89.2 | 70.9 | 77.1 | 77.2 | 63.0 |
| Age 3 | 14.9 | 35.0 | 30.9 | 30.8 | 23.3 | 19.2 | 16.9 | 17.3 | 15.6 |
| Fishable stock | 1916 | 2352 | 2019 | 2088 | 1885 | 1584 | 1620 | 1642 | 1424 |
| Calculated TAC | 1200 | 1635 | 1265 | 1420 | 1285 | 975 | 1050 | 1040 | 835 |
| Advised TAC | 1390 | 1600 | 1265 | 1200 | 1000 | $* * 1090$ | 1350 | 1005 | 875 |

*In January 1993, 80000 t were added to the 820000 t recommended after the October 1992 survey due to an unexpected large increase in mean weights.
** In March 2001, 100000 t were added to the 990000 t recommended after the January/February 2001 survey due to much higher mean weights in the catch during 1 February-10 March than measured during the survey.

Table 5.7.1 Capelin in the Iceland-East Greenland-Jan Mayen area 1978-2004. Recruitment of 1 year old fish (unit $10^{9}$ ) and total stock biomass (' 000 t ) are given for 1 August Spawning stock biomass (' 000 t ) is given at the time of spawning (March next year). Landings (' 000 t ) are the sum of the total landings in the season starting in the summer/autumn of the year indicated and ending in March of the following year.

| Year | Recruitment | Total <br> Stock biomass | Landings | Spawning <br> stock biomass |
| :--- | ---: | ---: | ---: | ---: |
| 1978 | 164 | 2832 | 1195 | 600 |
| 1979 | 60 | 2135 | 980 | 300 |
| 1980 | 66 | 1130 | 684 | 170 |
| 1981 | 49 | 1038 | 626 | 140 |
| 1982 | 146 | 1020 | 0 | 260 |
| 1983 | 124 | 2070 | 573 | 440 |
| 1984 | 251 | 2427 | 897 | 460 |
| 1985 | 99 | 2811 | 1312 | 460 |
| 1986 | 156 | 3106 | 1333 | 420 |
| 1987 | 144 | 2639 | 1116 | 400 |
| 1988 | 81 | 2101 | 1037 | 440 |
| 1989 | 64 | 1482 | 808 | 115 |
| 1990 | 118 | 1293 | 314 | 330 |
| 1991 | 133 | 1675 | 677 | 475 |
| 1992 | 143 | 2058 | 788 | 499 |
| 1993 | 144 | 2287 | 1179 | 460 |
| 1994 | 224 | 2287 | 864 | 420 |
| 1995 | 197 | 3007 | 929 | 830 |
| 1996 | 191 | 2885 | 1571 | 430 |
| 1997 | 165 | 2348 | 1245 | 492 |
| 1998 | 168 | 2197 | 1100 | 500 |
| 1999 | 138 | 2315 | 933 | 650 |
| 2000 | 146 | 140 | 2164 | 1071 |

*Preliminary
** Adults only
NA: Not available

Figure 5.5.2.1. The relationship between the measured numbers of immature 1 -group capelin in autumn acoustic surveys and the numbers of maturing capelin in 1 August of the following year (left hand figure) and between measuredtotal numbers of 2-group capelin and the maturing 3-group capelin in the following year (right hand figure).


Figure 5.5.2.2. The relationship between the total numbers in the maturing stock and the mean weight of maturing 2group (left hand figure) and 3-group (right hand figure) in autumn 1989 ö2001.


## BLUE WHITING

Blue whiting is a pelagic gadoid which is widely distributed in the eastern North Atlantic. The highest concentrations are found along the ridge of the continental shelf in areas II, V, VI at depths ranging between 300 and 600 meter but is also present in almost all other management areas between the Barents Sea and the Strait of Gibraltar. Morphological, physiological, and genetic research has suggested that there may be several components of the stock which mix in the spawning area west of the British Isles (ICES C.M. 2000/ACFM:16; Heino et al. 2003). For assessment purposes blue whiting in these areas is treated as a single stock since it has so far not been possible to define an unambiguous border between populations.

### 6.1.1 ACFM advice and management applicable to 2003 and 2004

In 1998 ACFM defined limit and precautionary reference points for this stock: $\mathbf{B}_{\text {lim }}$ ( 1.5 mill. t.), $\mathbf{B}_{\mathrm{pa}}$ ( $2.25 \mathrm{mill}$. t.), $\mathbf{F}_{\text {lim }}$ ( 0.51 ) and $\mathbf{F}_{\mathrm{pa}}$ (0.32). The advice of ACFM in following years has been given within a framework defined by these reference points.

In 2002 ACFM stated that "the stock is harvested outside safe biological limits. The spawning stock biomass for 2001 at the spawning time (April) is inside safe biological limits while the SSB for 2002 is expected to be below $\mathbf{B}_{\mathrm{pa}}$. Fishing mortality has increased rapidly in recent years, and is estimated at 0.82 in 2001. Total landings in 2001 were almost 1.8 million t . The incoming year classes seem to be strong. ICES recommends that the fishing mortality be less than $\mathbf{F}_{\mathrm{pa}}=0.32$, corresponding to landings of less than 600000 t in 2003. Implementation of a rebuilding plan is not necessary since according to the current assessment the state of the stock is better than previously estimated".

In 2003 ACFM stated that "the current estimates of SSB and fishing mortality are uncertain. Nevertheless, the spawning stock biomass in 2003 is likely to be above $\mathbf{B}_{\mathrm{pa}}$. Therefore, based on the most recent estimates of fishing mortality and SSB, ICES classifies the stock as likely to be harvested outside safe biological limits ( $\mathrm{F}>\mathbf{F}_{\text {lim }}$ ). Total landings in 2002 were almost 1.6 million t . The incoming year classes seem to be strong. ICES recommends that catches should be less than 925000 tonnes in 2004 in order to achieve a $50 \%$ probability that the fishing mortality in 2004 is less than $\mathbf{F}_{\mathrm{pa}}$ $(=0.32)$. This will also assure a high probability that the spawning stock biomass in 2005 will be above $\mathbf{B}_{\mathrm{pa}}{ }^{\prime}$.

In December 2002 EU, Faroe Islands, Iceland, and Norway agreed to implement a long-term management plan for the fisheries of the blue whiting stock, which is consistent with a precautionary approach, aimed at constraining the harvest within safe biological limits and designed to provide for sustainable fisheries and a greater potential yield. The plan should consist of the following:

1. Every effort shall be made to prevent the stock from falling below the minimum level of Spawning Stock Biomass (SSB) of 1500000 tonnes.
2. For 2003 and subsequent years, the Parties agreed to restrict their fishing on the basis of a TAC consistent with a fishing mortality less than 0.32 for appropriate age groups as defined by ICES, unless future scientific advice requires modification of the fishing mortality rate.
3. Should the SSB fall below a reference point of 2250000 tonnes ( $\boldsymbol{B}_{p a}$ ) the fishing mortality rate, referred to under paragraph 1, shall be adapted in the light of scientific estimates of the conditions then prevailing. Such adaptation shall ensure a safe and rapid recovery of the SSB to a level in excess of 2250000 tonnes.
4. In order to enhance the potential yield, the Parties shall implement appropriate measures, which will reduce catches of juvenile blue whiting.
5. The Parties shall, as appropriate, review and revise these management measures and strategies on the basis of any new advice provided by ICES.

The agreed management plan was, however, not implemented. Although the Coastal States and Russia were able to agree on this management plan for blue whiting in 2002, in the same vein as earlier agreements on mackerel and herring, it has not been possible to reach an agreement on the allocation of blue whiting and this has been a major outstanding issue for NEAFC in recent years. All parties want as large a percentage as possible and the dispute about percentages has dragged on. In the absence of agreements on a TAC for 2002 and 2003 and the allocation of the TAC, the Coastal States and the Russian Federation implemented unilateral measures to limit blue whiting catches for these
years. TAC's have been set by EU and Iceland only. During the year 2003, the EU increased its TAC for international waters with 250000 t . During 2004 the year 2004 it increased its total TAC with 350000 t applicable to all areas. The fisheries by Norway, Russia and Far Oer are not restricted by TACs.

### 6.2 Fisheries in 2003

Total catches figures in 2003 were provided by members of the WG. They were estimated to be 2.3 million t compared to 1.6 million t last year. Time series with catches by nations and area are given in Tables 6.2.1-6.2-7.

Spatial and temporal distribution of the catches of blue whiting in 2003 is given by quarter and ICES rectangles in Figure 6.2.1. The distribution of the catch by ICES rectangles for the whole year is given in Figure 6.2.2. In 2003 the catch provided as catch by rectangle represented approximately $99.9 \%$ of the total WG catch.

### 6.2.1 Description of the national fisheries

## Denmark:

The Danish directed fishery blue whiting fishery is mainly conducted by trawlers using a minimum mesh size of 40 mm . Blue whiting is also taken as by-catch using trawl with mesh sizes between 16 and 36 mm for Norway pout. This fishery is minor. The directed fishery caught $74,400 \mathrm{t}$ mainly in Divisions IIa, IVa, Vb1 and VIa with small catches from Divisions IIIa and XII. By-catches of blue whiting ( $8,500 \mathrm{t}$ ) are caught mainly in the Norway pout fishery in the North Sea and in the Skagerrak. Some blue whiting by-catches are also taken during the human consumption herring fishery in the Skagerrak.

## Germany:

The main German fleet fishing for pelagic species is based in Bremerhaven and Rostock. It consists of 4 large pelagic freezer trawlers with overall lengths of $70 \mathrm{~m}, 95 \mathrm{~m}, 117 \mathrm{~m}$ and 124 m ; engine power $3700 \mathrm{hp}, 4400 \mathrm{hp}, 8300 \mathrm{hp}$ and 8600 hp and $25-40$ crew members. The vessels are owned and managed by Dutch companies but operating under German flag.

They are specially designed for pelagic fisheries: The catch is pumped into large refrigerated seawater tanks to keep the catch fresh until further processing, which is usually block freezing. The freezing capacity of the 4 vessels ranges between 60,000 and 230,000 cartons (up to 5,500 t fish).

The German Blue Whiting fishery is conducted in the first and second quarter. The fishery starts in Divisions VIIb, c and XII in March and continues through Divisions VIa and bin April to divisions Vb1 and 2 in May. Occasionally there is a minor blue whiting by-catch in the directed herring fishery in Div. IIa and b in quarter 3. Very little discards have been reported by observers from the mackerel and horse mackerel fishery in the $4^{\text {th }}$ quarter in area VII.

## Faroe Islands:

In January the fishery ( 9 combined purse seiners/trawlers) concentrated on the southern part of the Faroese EEZ (ICES Division Vb). In February the fishery had moved south to the Porcupine area west of Ireland (Division VIIc,k and VIb) in the EU zone. Later in February and March the fishery continued in international waters Div. VIb, VIIc, k and XII outside the EU zone with a few catches in the EU zone. In early April the fishery continued in international waters (VIb), however, most catches were taken in the northern part of VIb (EU zone) and in the southern part of the Faroe zone $(\mathrm{Vb})$. In May the fishery continued in the southern part of the Faroese area and on the slope vest off the Faroe plateau, indicating that the fish migrated west of the Faroes on their way northwards. Later in May the fishery moved northwest towards the Icelandic border and into the Icelandic zone (Divisions Va and Vb ). In June the fishery operated in the Faroese and Icelandic zones (Va and IIa) and continued in this area in July and August, also including Division Vb . The fishery continued on the Faroe-Iceland ridge in September and October, gradually moving closer towards the Faroe Islands, and in November till the end of 2003 the catches were exclusively taken in Faroese waters (IIa and Vb ). More than $90 \%$ of the catches were taken with pelagic trawl ( $44 \mathrm{~mm} \#$ in cod end).

In August some blue whiting was caught as by-catch in the herring fishery in the Svalbard zone. The industrial fleet (3 trawlers) operated mainly in Norwegian waters (ICES Division IVa) in 2003 with catches of blue whiting scattered throughout the year.

## France:

No information available.

## Iceland:

Iceland and Faroes have a bilateral agreement of mutual fishing rights for blue whiting in each other's EEZs. Iceland set a total blue whiting catch quota of 547000 tonnes in 2003 for the Icelandic- Faroese EEZs and international waters

The Icelandic directed fishery for Blue whiting started on 11 March in International waters west of the British Isles (ICES Divisions XII, VIb) and about the same time in Icelandic waters with small catches at SE-Iceland. In April the main fishery had moved further north and was largely conducted in Faroese waters SW of the Faroes but with some small catches in International waters west of the British Isles and in Icelandic waters.

In early May the fishery was mainly conducted in the Faroese zone but shifted gradually into the Icelandic zone in the latter half of May. In June, July and August most of the Icelandic Blue whiting catch was fished in the Icelandic area but in September the fishery moved increasingly into the Faeroese zone. In October there was again an increase in catches in the Icelandic area with more than half the catch taken there. In November and December most of the fishery took place in the Faroese zone. The total Icelandic catch was 501494 tonnes.

Iceland has set size limitations on landings of blue whiting. If the catch consists of $30 \%$ or more of fish smaller than 25 cm , a temporary area closure is imposed.

A total of 21 purse-seiners/trawlers participated in the Icelandic fishery. The length range of the vessels was 51-86m with a mean length of 66 m . The engine power range of the fleet was 1943-5520 kW ( $2640-7500 \mathrm{HP}$ ) with a mean of 3593 kW ( 4882 HP ). The total average catch/hour for individual vessels was in the range of 6.9-36.3 tonnes with an average of 17.6 tonnes for the whole fleet.

## Ireland:

There have been Irish fisheries for blue whiting since the late 1980's, when larger recirculated sea water (RSW) vessels carried out exploratory voyages. However the current Irish fishery for blue whiting developed in response to restrictive quotas for mackerel and herring in the 1990's. Catches peaked in 1998, when approximately $46,000 \mathrm{t}$ were landed by 7 vessels. These vessels are the larger Irish RSW and freezer trawlers, in excess of 60 m length. In 2003, the total Irish catch was $22,580 \mathrm{t}$, landed by 6 vessels. In 2003, the EU TAC for Sub-areas V, VI, VII, XII and XII was 107,281 t, with an Irish quota share of $17,165 \mathrm{t}$. However additional allocations of quota were made to Ireland in 2003.

Fishing takes place from March to May between the Porcupine (VIIbcjk) and Rockall Banks (VIb) using large single trawls that have been specially modified to take bulk catches from deep water. Maximum circumference of the gear is $1,700 \mathrm{~m}$ with a mesh-size in the cod-end liner of 35 to 40 m . The Irish fishery is dominated by landings for fish meal, but there is some activity by freezer trawlers, for human consumption.

## Netherlands:

The Dutch fleet fishing for pelagic species (herring, blue whiting, mackerel, horse mackerel, argentines) consists of 16 freezer trawlers and pair trawlers. The fishery directed towards blue whiting takes place during the $1^{\text {st }}$ half of the year, mainly in ICES Division Vb, VIa,b and VIIb,c. Total catches in 2003 were 48000 t and are landed frozen for human consumption. There is a seasonal fishery in May-June directed towards NSSH in international waters (ICES Sub-area II) during which they also catch some blue whiting.

## Norway:

Norway set a blue whiting quota of 250000 t for the directed fisheries in the Norwegian EEZ, Jan Mayen zone and international waters for 2003. In addition, through international agreements, 120000 t in the EEZ of EU and 36200 t in the Faroese zone were made available to the Norwegian fleet. A separate quota of 80000 t was given to the mixed industrial fishery in the Norwegian EEZ in the North and Norwegian Seas south of $64^{\circ} \mathrm{N}$. The total quota for Norwegian vessels in 2003 was 486200 t .

The main Norwegian fishery is a directed pelagic trawl fishery, regulated by vessel quotas, and is carried out on and west of the spawning areas west of the British Isles. A total of 46 large combined purse seiners/trawlers took part in the fishery in 2003.

In 2003, as usual, there was a seasonal progression of the fishery from the international waters off Porcupine Bank and Rockall in the beginning of the season (January-March) towards the shelf edge in EU zone and the banks in the Faroese waters in the end of season. The fishery in EEZ of EU was stopped 30 April and that in the Faroese zone 2 May after the quotas in the respective zones were taken. In contrast to the usual pattern, a directed fishery continued in the Norwegian EEZ where significant proportions of young blue whiting, particularly of age 1 years, occurred in the catches. The fishery in the Norwegian EEZ, Jan Mayen zone and international waters was stopped on 30 June. At this point, the quota was exceeded by $115 \%$. The total catch in the directed fishery was about 703000 t , which is a new record.

Blue whiting were also fished in the North Sea and in the southern Norwegian Sea (areas south of $64^{\circ} \mathrm{N}$ ) in the mixed industrial fishery targeting blue whiting and Norway pout. The quota regulation was not put in effect, and an estimated catch of approximately 131000 t was taken in this fishery in 2003; this a new record.

North of $64^{\circ} \mathrm{N}$ in area IIa, small amounts (about 1336 t ) of blue whiting were caught as by-catch and landed by vessels targeting argentines along the shelf edge. These catches are included in the numbers for the mixed industrial fishery.

## Portugal:

In 2003 Portuguese landings of blue whiting amounted to 2652 tonnes. Most of the catch was taken by a trawl fishery ( $87 \%$ ) The rest was taken by artisanal ( $13 \%$ ) and purse seine ( $0.2 \%$ ) fisheries. About $68 \%$ of the blue whiting was landed by trawlers in two main harbours, Portimão, on the southern coast and Matosinhos on the northern coast. The rest of the landings took place in other 15 harbours.

## Russia:

At the beginning of the year 10-15 fishing vessels operated in the blue whiting fishery in the south of the Norwegian Sea. From 8 February fishery to the southwest of Rockall started. At the beginning of March 9 vessels operated there, and in the end of that month their number increased to 14 . Spawning blue whiting was distributed at depth between 450 and 600 m . In the first half of April the catch rates were declining with fish migrating to the northeast. By mid-April the fishery at Rockall practically finished. It started there 2 weeks earlier and finished one week later than in 2002. The vessels moved to the southern part of the Norwegian Sea. Catch rates there were increasing until about mid-April, when they declined slightly, however, remained fairly good for directed fishery until mid-June. The first trawlers came to the central part of the Norwegian Sea at the beginning of May, and their number was growing with increasing catch rates. However, already in July no more than 5 fishing vessels were operating there targeting blue whiting. The scale of this fishery increased again in the end of October, when up to 30 vessels engaged in the fishery in the north of the Norwegian Sea. The fish remained there longer than in 2002, when as early as in the middle of the month it began to follow the southward migrations of the fish. In 2003 more westerly migration pattern was observed, which forced the vessels to stop fishery in mid-November. Some of these vessels moved to fishing blue whiting in the southern part of the sea.

## Spain:

The Spanish blue whiting fishery was carried out mainly by bottom pair trawlers in a directed fishery and by single bottom trawlers in a by-catch fishery, and small quantities ( 42 t ) were also caught by longliners. These coastal fisheries have trip durations of 1 or 2 days and catches are for human consumption. Thus, coastal landings are rather stable due mainly to market forces.

Pair Bottom Trawl Fishery: The Pair bottom trawl is a traditional fleet that fish mainly blue whiting (above 80\%) and other pelagic species in Div. VIIIc and North IXa. In the middle of 90 's, VHVO gear (with 25 m of vertical opening) was gradually substituting the traditional one. From 2001 the cod end mesh size was increased to 55 mm . This fleet is composed of 68 pairs ( 136 vessels) with an average of 25 m length, 472 HP , and 145 GRT. The pair trawler fleet landed 12550 t , taken mainly on the border between Divisions VIIIc and IXa.

Bottom Trawl Mixed Fishery: This metier operates in Divisions VIIIc and IXa North, using a cod end mesh size of 65 mm and a vertical opening of 1.2-1.5 m . It targets a wide range of species including horse mackerel, blue whiting, mackerel, hake, anglerfish, megrims, and Nephrops (the first three species contributes around $70 \%$ of the landings). At
present it is composed of around 235 vessels, with an average of 27 m length, 561 HP , and 177 GRT. By-catches of blue whiting in this fishery were ( 1233 t ).

The decrease in the Spanish landings is partly explained by the closure of the fishery in some fishing grounds and ports, due to the sinking of the oil tanker "Prestige" in Galician waters in December 2002. These closures took place in the first semester of 2003.

## Sweden:

The fishery for blue whiting did expand substantially during 2003. It is pursued by pelagic trawlers (both single and pair-trawlers). Almost $80 \%$ is landed in Norway, the rest is landed in Denmark, Iceland, Faroes, United Kingdom and Sweden. Nearly all the catch is taken in the 3rd quarter.

## UK (Scotland):

The Scottish directed fishery for blue whiting takes place during the first and second quarters of each year. Approximately one third of the pelagic fleet participates in this fishery which is pursued mainly in ICES areas VI and VII. The vessels operating in this fishery are large pelagic trawlers using industrial fishing gear. Generally the catch is landed for fish meal, into factories in Shetland, Faroes, Ireland and Denmark and is worth around $£ 2$ million to the Scottish fleet. A summary of the fleet structure is given in Tables 3.1.3.1 and 6.2.8.

### 6.3 Biological Characteristics

### 6.3.1 Length composition of Catches

Data on the combined length composition of the 2003 commercial catch by quarter of the year from the directed fisheries in the Norwegian Sea and from the stock's main spawning area were provided by Denmark, the Faroes, Iceland, Ireland, the Netherlands, Norway and Russia. Length composition of blue whiting varied from 12 to 44 cm , with $67 \%$ of fish ranging from $23-28 \mathrm{~cm}$ in length and $98 \%$ from $18-33 \mathrm{~cm}$. The mean length in this fishery was 25.1 cm (Table 6.3.1.1). Length compositions of the blue whiting catch and by-catch from "other fisheries" in the Norwegian Sea (south of $64^{\circ} \mathrm{N}$ ), the North Sea and Skagerrak were presented by Norway, Denmark and Sweden (Table 6.3.1.2). The catches of blue whiting from the mixed industrial fisheries consisted of fish with lengths of $12-37 \mathrm{~cm}$ and a mean of 21.2 cm . Spain and Portugal caught blue whiting in the Southern area. The Spanish data used for length distribution of catches showed a length range from $11-40 \mathrm{~cm}$ with a mean length of 21.1 cm (Table 6.3.1.3).

### 6.3.2 Age composition of catches

For the directed fisheries in the northern area in 2003, age compositions were provided by Denmark, the Faroe Islands, Germany, Iceland, Ireland, Norway, the Netherlands and Russia, and the sampled catch accounted for $89 \%$ of the total catch. Estimates of catch in numbers for unsampled catches were raised according to the knowledge of how, where, and when the catches were taken. The age compositions in the directed fisheries are given in Table 6.3.2.1.

Age compositions for blue whiting by-catches from "other fisheries" in the North Sea and Skagerrak were provided by Norway and sampled catch accounted for $82 \%$ of catches. These data were used for allocation of the remaining part of the total in that area. The age compositions are given in Table 6.3.2.2.

For the fisheries in the Southern area, age compositions representing $100 \%$ of the catch were presented by Spain and Portugal. The age compositions in the southern fishery are given in Table 6.3.2.3.

The combined age composition for the directed fisheries in the Northern area, i.e. the spawning area and the Norwegian Sea, as well as for the by-catch of blue whiting in "other fisheries" and for landings in the Southern area, were assumed to represent the overall age composition of the total landings for the blue whiting stock. The catch numbers-at-age used in the stock assessment are given in Table 6.3.2.4. The 2000 year class was the most numerous in the catches, followed by the 1999, 2001 and 2002 year classes. To calculate the total international catch-at-age, and to document how it was done, the program SALLOC was used (ICES 1998/ACFM:18). The allocations are shown in the Annex II.

Mean weight-at-age data were available from Denmark, the Faroes, Germany, Iceland, Ireland, the Netherlands, Norway, Portugal, Russia, and Spain. Mean weight-at-age for other countries was based on the allocations shown in the Annex II ("ALLOC" files) and was estimated by the SALLOC program for the total international catch. Table 6.3.3.1 shows the mean weight-at-age for the total catch during 1981-2003 used in the stock assessment. The weight-at-age for the stock was assumed to be the same as the weight-at-age for the catch.

### 6.3.4 Maturity-at-age

Maturity-at-age used in the assessment was obtained by combining maturity ogives from the southern and northern areas, weighted by catch in numbers-at-age (ICES 1995/Assess:7). These are the same as those used since 1994 (Table 6.3.4.1). Although the values of maturity-at-age probably are too low, sufficient information for estimating new ogives is not available.

### 6.3.5 Natural Mortality

The possible need for revising the current estimate of instantaneous natural mortality rate $M$ for blue whiting was discussed in detail by the 2002 WG . Although it was admitted that the current estimate $M=0.2 \mathrm{yr}^{-1} \mathrm{might}$ be too low, the factual basis for revision was ambiguous. More recent methodological work by WGMG (ICES 2003/D:03) emphasises that natural mortality rate cannot be estimated reliably with information normally available for stock assessment models. WG therefore considers that there is no new information that would justify a revision of the current estimate of $M$.

### 6.4 Stock estimates

### 6.4.1 Acoustic surveys

The time series from the acoustic surveys used in the analytic assessment are given in Tables 6.4.1.1-6.4.1.3.

### 6.4.1.1 Surveys in the spawning season

An international blue whiting spawning stock survey was carried out on the spawning grounds west of the British Isles in March-April 2004 (Heino et al., WD a). Four vessels participated in the survey: "Celtic Explorer", "Fridtjof Nansen", "Johan Hjort" and "Tridens". This is the first international survey with such a broad international participation, which allowed for broad spatial coverage as well as a relatively dense net of trawl and hydrographical stations.

The highest abundances of blue whiting were observed along the shelf edge from the northern Porcupine bank to the Hebrides, south/southwest of the Faroe Plateau and the adjacent banks, and west of Rockall. Limits of the oceanic distribution were found in south and southwest. The distribution of acoustic backscattering densities for blue whiting as recorded by the four vessels are shown in Figure 6.4.1.1.1.

The blue whiting spawning stock estimate based on the international survey is 10.6 million tonnes and $119 \times 10^{9}$ individuals. An age-disaggregated total stock estimate is presented in Table 6.4.1.1.1, showing that the stock was dominated by blue whiting of 4 years in age (year class 2000). Age class 3 years was also abundant, followed by age classes 2 and 5 years (Figure 6.4.1.1.2). Blue whiting of ages $3-5$ years made up $78 \%$ of spawning stock biomass. There was some variability in the age structure between different areas with the highest mean age observed in the Hebrides area.

Results from the individual vessels are reported in more detail in respective survey reports (O'Donnell et al., WD; Heino et al., WD b). Of these the Norwegian one is used in tuning the assessment. For this reason, it was decided to keep the spatial coverage of this survey similar to earlier years although a more systematic survey design was adopted (ICES 2003/D:10). As shown in the text table below, the results from this survey are very much in agreement with those from the international survey: the spawning stock is estimated to amount to 10.9 million tonnes and $128 \times 10^{9}$ individuals; this is in accordance with a slightly smaller average size recorded in the Norwegian in comparison to the international survey. There is also a slight difference in age distribution in favour of younger fish in the Norwegian survey.

| Survey | $\begin{array}{c}\text { Abundance, } 10^{9} \\ \text { individuals } \\ \text { total }\end{array}$ |  | spawning |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | Biomass, mill. tonnes \(\left.\left.\begin{array}{c}total <br>

spawning\end{array} $$
\begin{array}{c}\text { Mean } \\
\text { weight, } \\
\mathrm{g}\end{array}
$$\right) \begin{array}{c}Mean <br>
length, <br>

\mathrm{cm}\end{array}\right]\)| International <br> (incl. Norway) | 130 | 119 | 11.1 | 10.6 | 85.5 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Norwegian | 137 | 128 | 11.4 | 10.9 | 83.2 |

In comparison to earlier years, both total and spawning stock biomass estimates are similar to those obtained in 2002 and 2003. The stock in 2004 is, however, composed of fewer and larger fish than in 2002 and 2003. To a large extent this reflects the ageing of year class 2000 that has numerically dominated the total stock estimate since 2002. In fact, this year class has broken the earlier record numbers each year in 2002-2004.

An age-stratified stock estimate from the Norwegian survey is given in Table 6.4.1.1.2. This shows the dominance of blue whiting of age 4 years, but that also age group 3 years is relatively abundant. In comparison to earlier years, absolute numbers of blue whiting of age 2 and 3 years are lower than estimated in 2002 and 2003, but those of age 5 years are more abundant. Age and length distributions are displayed in Figure 6.4.1.1.4.

Estimates of total and spawning biomass of blue whiting in the spawning area based on Russian and Norwegian surveys since 1983, together with the one from the international survey in 2004, are given in Table 6.4.1.1.3. Usually, acoustic estimates have been well above the analytical assessments. Therefore the acoustic estimates have been used only as relative indices. A factor contributing to the high acoustic estimates is the acoustic target strength that is probably too low.

### 6.4.1.2 Surveys in the feeding area

Since 1995, Norway, Russia, Iceland, and Faroes, and in the period 1997-2000 the EU, have coordinated their survey efforts on pelagic fish stocks in the Norwegian Sea. In 2003 these were coordinated through PGSPFN, and a detailed description of the results can be found in the corresponding report (ICES 2003/D:10). Here the three surveys where blue whiting were among the focal species are briefly reviewed.

An international acoustic survey, with one vessel from the Faroes, Iceland and Norway each, covered most of the temperate parts of the Norwegian Sea in May-June 2004. This survey covers post-spawning adult blue whiting in the south-western part of the survey area as well as juvenile and adult fish in the Norwegian Sea. Total stock estimate from the survey was almost twice as high as in the previous year (see the text table below). However, this increase was largely due to more extensive coverage in the south-western parts of the survey area. Results from comparable surveys from 2000 onwards are presented in Table 6.4.1.2.1; these are calculated for the area north of $63^{\circ} \mathrm{N}$ and east of $8^{\circ} \mathrm{W}$ that has had good annual coverage during all these years (see Heino et al., WD c,d). The time series shows a relatively low total stock biomass in the survey area in 2000, and much higher and relatively stable biomass in 2001-2003. In 2003 year classes 2000 and 2002 were most numerous. The stock biomass was dominated by year class 2000 .

Iceland has carried out an acoustic survey in summer in the Icelandic EEZ since 1998; age-disaggregated results are available from 1999 onwards (Table 6.4.1.2.2). The survey in 2003 gave stock estimate that was well above the one from the previous year (see the text table below). However, the survey in 2003 had a wider coverage than those before. This reflects wider distribution of blue whiting in the Icelandic waters, and it could be argued that the increase reflects genuine increase of blue whiting in the area (see the discussion in Heino et al., WD c,d). Table 6.4.1.2.2 gives nevertheless also an estimate that is calculated for the more restricted area covered all years. The results show rather broad age structure. Year class 2002 is the most numerous one. In the wider survey area, year class 2000 is also very abundant. Virtually all 0 -group blue whiting were observed in the core survey area and had a relatively low abundance.

|  | Total stock biomass (million t$)$ |  |
| :---: | :---: | :---: |
| Year | International survey in May-June | Icelandic survey in June-August |
|  |  | 6.6 |
| 1998 | 4.2 | 1.6 |
| 1999 | 2.5 | 1.8 |
| 2000 | 5.9 | 1.2 |
| 2001 | 6.0 | 2.1 |
| 2002 | 11.7 | 1.9 |
| 2003 |  | 3.1 |

Norway conducted a survey in the northern Norwegian Sea in July-August (ICES 2003/D:10). Although this survey is similar in timing to the survey that is used as a tuning time series (Table 6.4.1.3), the coverage in 2003 (and in 2002) was much reduced in comparison to the earlier years. The survey is therefore not comparable to the one in 2001 and before. In comparison to 2002, the survey suggests small decreases in biomass ( $-13 \%$ ) and stock numbers $(-21 \%)$. The most abundant year class was that from 2000.

### 6.4.1.3 Russian survey in the Barents Sea

The Russian research vessels tried to assess biomass of the blue whiting in the Barents Sea in 2000 for the first time. In October-December 2003 three research vessels in the Barents Sea and adjacent waters conducted multi-species trawlacoustic survey on demersal fish including blue whiting.

Estimated biomass of blue whiting over the whole surveyed area constituted nearly 350 thousand tonnes and abundance was 5,3 milliard individuals. The maximum quantity of fish estimated to be about 250 thousand tonnes occurred in the ICES Div. 11a.

Length frequencies of blue whiting varied from 18 to 41 cm with individuals oft $24-28 \mathrm{~cm}$ long being predominant.

### 6.4.2 Bottom trawl surveys

### 6.4.2.1 Surveys in the southern areas

Bottom trawl surveys have been conducted off the Galician (NW Spain) and Portuguese coasts since 1980 and 1979 respectively, following a stratified random sampling design and covering depths down to 500 m . Since 1983, the area covered in the Spanish survey was extended to completely cover Spanish waters in Division VIIIc. The area covered in the Portuguese survey was also extended in 1989 to the 750 m contour. A new stratification in the Spanish surveys has been established since 1997. Stratified mean catches and standard errors from the Spanish and Portuguese surveys are shown in Tables 6.4.2.1 and 6.4.2.2. In both areas larger mean catch rates are observed in the $100-500 \mathrm{~m}$ depth range. Since 1988 the highest catch rates in the Spanish survey were observed in 1999 ( $124 \mathrm{~kg} / \mathrm{haul}$ ). The 2003 estimate is relatively low in both the Spanish ( $58 \mathrm{~kg} /$ haul) and the Portuguese autumn survey ( $76 \mathrm{~kg} / \mathrm{haul}$ ). The Portuguese autumn surveys generally give higher values than in the summer surveys, and a better correlation with the Spanish surveys (Figure 6.4.2.1).

### 6.4.2.2 Surveys on the Faroe plateau

On the Faroe plateau two annual demersal bottom trawl surveys are carried out, the spring survey in March (1996-2003) and the autumn survey in August/September (1994-2003). The surveys are aimed at cod, haddock and saithe, but varying amounts of blue whiting are caught as by-catch each year. The size of the blue whiting ranges between 9-45 cm during autumn, including 0 -group and older blue whiting, and between $11-45 \mathrm{~cm}$ during spring, including 1 -group and older fish.

After the spawning west of the British Isles and Ireland in spring the larvae drift northwards with the currents passing the Faroes and young of the year are found in the deeper regions of the shelf in August. 0-group blue whiting are also found on the shelf in the deeper regions and are caught in both the autumn and spring groundfish surveys.

Unfortunately no otoliths are available from these groundfish surveys prior to 2004, making the separation of the 0 - and 1 -group blue whiting inaccurate. For the summer survey Icelandic age-length keys (from June-August) were used to split the length into age groups (Table 6.4.2.3). Since the Icelandic age-length keys were from the June-August period, the cannot be used to split the spring data for two reasons, firstly because in spring the 0 -group is not present in the catches, whereas is present in the June-August age-length keys, and secondly because the length growth is very fast at the young stages of blue whiting and the resulting age distribution in spring would be biased. Therefore the 1 -group were visually separated from the length distribution from the spring surveys (Table 6.4.2.4). In the spring surveys the 1group is usually well separated in the length distribution (Table 6.4.2.5). In 2004 the 1 -group was separated from otolith samples during the survey. Blue whiting otoliths will be sampled in future bottom trawl surveys.

The consistency in the time series in representing the 0 - and 1 -group can be checked by a comparison of the abundance of the 0 -group caught in summer with the abundance of 1 -group the following spring. Figure 6.4 .2 .2 shows the bottom
trawl 0-group index (number of blue whiting caught $10^{-3}$ ) from autumn 1996-2003 compared to the following 1-group spring index lagged one year to match the 0 -group index, the correlation is very high, the $\mathrm{r}^{2}$ is around 0.98 .

### 6.4.2 3 Survey in the North Sea

Norway has conducted a bottom trawl survey on shrimp in the northern parts of the North Sea annually in October since 1984 (Tveite 2000). Blue whiting is caught in $>95 \%$ of trawl hauls, with individual lengths ranging between 9 and 52 cm . Blue whiting have not been aged, but based on sampling of the commercial catches in the area by Norway, a reasonably good separation of 0 -group blue whiting can be achieved by assuming that all individuals less than 19 cm in body length are of age 0 years. Separation of other age groups have not been attempted. The results are summarized in Table 6.4.2.6 and show that 0 -group blue whiting are occasionally very abundant, but that the years of great abundance show only weak correspondence with large year classes seen in the analytic assessment. This applies in particular to the recent strong year classes that have not occurred in large numbers in the North Sea.

### 6.4.2 $4 \quad$ Surveys in the Barents Sea

Norway has conducted bottom trawl surveys targeting cod and other demersal fish in the Barents Sea since late 1970s. From 1981 onwards there have been systematically designed surveys carried out during the winter months by at least two Norwegian vessels; in some years the survey has been conducted in co-operation with Russia. Blue whiting is a regular by-catch species in these surveys, and is in some years one of the most abundant species. Most of the blue whiting catches (or samples thereof) have been measured for body length, but very few age readings are available (starting from 2004 otoliths will be collected systematically). The existing age readings suggest that virtually all blue whiting less than 21 cm in length belong to 1 -group and that while some 1 -group blue whiting are larger, the resulting underestimation is not very large. An abundance index of all blue whiting and supposed 1 -group blue whiting from 1981 onwards is given in Table 6.4.2.7 and follows methods described in Heino et al. (2003b). The data suggest that there is good correspondence between 1-group blue whiting in the Barents Sea and estimated recruitment in the main Atlantic stock: correlation between the 1 -group index and recruitment as estimated in the final AMCI in 2004 assessment is 0.88 . Earlier analyses also suggest that large inFlows of warm Atlantic water tend to result in higher abundance of blue whiting in the area (Heino et al. 2003b). Whether this holds for 1-group blue whiting in particular remains to be seen.

### 6.4.3 Catch per unit of effort

The Spanish pair trawl CPUE series (Table 6.4.3.1) has been used for several years as a tuning fleet in the blue whiting assessment. This fleet represents only a small part of the landings caught in a small part of the distribution area. Due to this fact, and following a recommendation of the Methods working group (ICES 2003/D:03), this tuning series is not used in the final assessments since 2003. The data show a slight decreasing trend in CPUE (Figure 6.4.3.1).

CPUE data in the spawning area is available from the Norwegian commercial fleet. The CPUE is calculated as an overall aggregated CPUE and is based on data from the vessels that have submitted their logbooks (in 2003 only 7 of 46 vessels did so, representing about $11 \%$ of catch). The data suggest an overall increasing trend in CPUE since early 1990s (Figure 6.4.3.2). However, no attempt has been made to correct the data for the increase in catchability caused by technological improvements, nor is a measure of variability available prior to 2003. As in previous years, the data have not been used in the assessment.

### 6.4.4 Data exploration

### 6.4.4.1 Analysis of catch-at-age in commercial data

Percentage distributions of catch of blue whiting by year class in 2003 by ICES Divisions grouped into larger areas are shown in Figure 6.4.4.1. The total catch at age is grouped into a northern area (I and IIa), central area (Iceland and Faroes; Va, b), North Sea (IVa and IIIa), the spawning area (VIa, b, VII and XII) and a southern area (VIII and IXa). The percentages of the catches in the area of the total catch is given below each chart.

The small 1-group immature fish (2002 year class) were mainly taken in the northern area (Norwegian Sea) and in the North Sea. Although high proportions of small fish were taken in the southern area, the contribution of these to the whole catch is insignificant. It should be noted that significant amounts of older blue whiting (ages 2-4) were also taken in the northern area. In the central area (Iceland and the Faroes) mainly $2-5$ year old blue whiting comprised the catches, or mainly mature fish. In the spawning area mainly 4 and 5 year olds were caught (ranging from 3-8 year olds) representing a third of the total catches.

### 6.4.4.2 Data exploration with AMCI

The AMCI assessment in 2003 was based on version 2.2, which at a later stage was found to contain a bug that affected the results. The bug was corrected in version 2.3, and that program version was used by SGAMHBW, who used artificial data to compare AMCI, ISVPA and SeaStar. Version 2.3 has thus been subject to certain quality control. Shortly before the WG meeting, a new version of AMCI was released, version 2.3a. The new version contains some additional features and bug fixes, but comparisons between versions 2.3 and 2.3 a did not reveal any differences, nor were the new features used. Comparison between version 2.2 and $2.3 / 2.3$ a with the settings from the final AMCI assessment in 2003 shows that recruitment is estimated similarly, but that the general level of spawning stock biomass is lower for the new versions and that general level of fishing mortality is higher for the new versions (Figure 6.4.4.2.1).

In order to evaluate robustness of the assessment, a set of assessment runs was conducted with the settings from the final AMCI assessment in 2003 as a starting point. The only difference in the settings was that recruitment was estimated up to the penultimate assessment year (in 2003 this value was fixed). The main diagnostic in these exploratory runs was the sensitivity of estimated values of spawning stock biomass, recruitment and fishing mortality to altered assumptions.

Catch-at-age data are input at yearly resolution (Table 6.3.2.4). However, AMCI operates internally on a quarterly basis. Similarly as in 2003 WG, The spawning stock is derived from the mean stock numbers in the first quarter, and the survey indices are related to the mean values in the survey season (Table 6.4.5.1.1). The yearly fishing mortality was split on quarters assuming that the proportion 0.35 of the total annual fishing mortality occurs in the first and in the second quarter, 0.2 in the third quarter, and 0.1 in the fourth quarter.

Estimation of recruitment in the final assessment year (final assessment year refers here to 2003 even though Norwegian spawning stock survey from 2004 is used in the assessment) is very fragile as there are only two data points that provide information on this: catch at age 1 years in 2003, and the Norwegian spawning stock survey 2004 estimate for age 2 years. The value of the former piece of information is restricted by variable selection on young fish, that of the latter by the partial recruitment to the spawning stock at that age. As the starting point, no estimation of recruitment in 2003 was attempted (see below for exceptions). Instead, information from surveys (summarized in section 6.4.4.6.2) was used to make a guesstimate, $20 \times 10^{9}$, a value similar to geometric mean recruitment in the period 1993-2002 in the final AMCI assessment.

In the final assessment from 2003, selection pattern in the first year was fixed to a pattern obtained from an earlier program run. This is a restrictive assumption, and it is not clear whether it was made intentionally. Nevertheless, estimating the selection affected only the earlier part of the assessment (Figure 6.4.4.2.2). Whether selection was assumed to equal over age range 8-10+ years of 6-10+ years had a negligible effect; the latter range was chosen for the final assessment as blue whiting is expected to be fully recruited to the fisheries by age 4 years, and residuals did not suggest any unwanted patterns. This is referred to as the "basic" model below.

Down-weighting of catch at age 1 year by a factor of 0.5 was applied by 2003 WG. Excluding this down-weighting results in some increase F and some decrease in recruitment and SSB in the most recent years (Figure 6.4.4.2.2).

SGAMHBW (ICES 2004/ACFM24) suggested excluding certain ages from the survey tuning time series (age 8 years from the Norwegian spawning stock survey, ages 7-8 years from the Russian spawning stock survey, and ages 5-7 years from the Norwegian Sea survey). This had little effect on the estimation results (Figure 6.4.4.2.2).

Eye-balling the catch at age data draws attention to very high catch numbers of blue whiting of age 1 years in 2003. In fact, the catch was second highest in record. This could happen if year class 2002 were particularly strong, or if selection pattern of the fisheries changed. Fisheries-independent survey data do not support the former alternative (see discussion above on determining recruitment in 2003), whereas a closer scrutiny of fisheries (section 6.4.4.1) supports the view that the selection pattern may have changed. In particular, Norwegian directed fisheries caught large numbers of young blue whiting in the southern parts of Norwegian Sea in a fishery that has no counterpart in the preceding ten years or so. This motivated a series of model runs to see the effect of this juvenile fishery on the assessment.

First, trying to estimate recruitment in 2003 was not successful (Figure 6.4.4.2.3): with the basic settings, this gives an estimate that is larger than recruitment in 1997 and 2001 (years with record recruitment in the final assessment), which is at odds with survey information. Including data from the Icelandic acoustic summer survey (ages 0-1 years) and/or the Barents Sea winter survey (age 1 years) did not make the situation better; rather, inclusion of the Barents Sea data caused convergence problems.

Second, with recruitment in 2003 assumed at $20 \times 10^{9}$ and "basic" settings as before, catch data corresponding to this age and year were given zero weight. This results in some increase in spawning stock biomass and recruitment and decrease in fishing mortality (Figure 6.4.4.2.3).

Finally, with recruitment in 2003 assumed at $20 \times 10^{9}$ and "basic" settings as before, gain factor at age 1 years was increased from 0.2 to 0.5 in order to let selection pattern to adjust better to fluctuations in fishing on juveniles. As in the previous case, this results in some increase in spawning stock biomass and recruitment and decrease in fishing mortality (Figure 6.4.4.2.3). On the basis of catch residuals (Figure 6.4.4.2.4) that showed less structuring for the later years than in the previous run, this model was chosen as the final AMCI model.

### 6.4.4.3 Data exploration with ISVPA

For blue whiting exploratory runs with ISVPA two versions of the model were carried out: 1) effort-controlled version, which give more confidence to the assumption of constant selection pattern in the fisheries Abundance-at-age is calculated via theoretical (model-derived) catches: 2) catch-controlled version, in which abundance-at-age is calculated directly via reported catch-at-age and the estimated selection pattern is used only for calculation of terminal abundances. But despite of "local" use of selection pattern in catch-controlled version, separabilization of the model gives the possibility to derive residuals for catch-at-age and to tell about solution determined only by catch at age data. Estimates produced by the catch-controlled version follow closer the observed catches and have lower residuals in catch-at-age, but are less resistant with respect to errors (especially outliers) in catch-at-age data.

As in previous years, in both versions of the model the parameter estimation procedure ensuring unbiased separable representation of fishing mortality was used.

In the assessments by both versions of ISVPA, three sets of age-structured survey data were used: Norwegian spawning acoustic, Russian spawning acoustic, and Norwegian sea acoustic. Each of the three survey data sets were split into two periods because of change in survey equipment which may have caused a change in catchability. The 2004 Norwegian spawning acoustic surveys was not included in the assessment, while the most recent year with data in Russian acoustic surveys is 1996, and 2001 in the Norwegian sea survey. That is why it is important to have additional signals from catch-at-age alone for determining the current status of the stock.

Exploratory runs with the catch controlled versions revealed that in comparison to the 2002 and 2003 assessments, when distinct minima were found in the SSQ or the median of distribution of squared residuals at a specific terminal fishing mortality in all the above mentioned sources of auxiliary data, in this year assessment only minima were found only in the second period of the Norwegian spawning acoustic survey and in the second period of the Norwegian sea surveys (see figures 6.4.4.3.1 and 6.4.4.3.2). However, the effort- controlled version gave more well-shaped minima, that is they give more distinct information about the stock.

The overall loss function, composed only from data sources, containing minima gives a more clear minimum in comparison to loss function composed of all available sources of information (compare two curves in the third column on figures 6.4.4.3.1 and 6.4.4.3.2). That is why for final ISVPA runs only data giving more clear signals about the current stock size were used (second period of Norwegian spawning acoustic, second part of Norwegian sea acoustic, and signal from catch-at-age). Incorporation of data giving no distinct minima will result in deterioration of the shape of overall signal and in driving the solution to improbably high and less certain values of fishing mortality in the terminal year.

It also should be mentioned that the signal about the current stock size from catch-at-age alone is weaker in comparison to the 2002 and 2003 assessment and required to use more robust measure of closeness: median of distribution of squared residuals, instead of their sum (see graphs of the first row in figure 6.4.4.3.1). This may be explained by a change in the catch-at-age data in 2003 or poorer quality of these data in comparison to previous years.

Application of the two versions of the model revealed very similar historical trends in SSB estimates (figure 6.4.4.3.3). Both versions showed an increase of SSB in 2003 with respect to 2002, but the effort-controlled version indicates that this increase is much lower in comparison to increase in SSB in recent previous years.

A comparison of modelled catches with the reported values for two versions of ISVPA are presented in Figure 6.4.4.3.4. The "catch-controlled" version of ISVPA follows more closely the reported catch values than the "effort-controlled" version of the program. It should be mentioned that for both versions, modelled catches in 2002 are higher than reported ones. This may indicate underreported catches in that year.

Figures 6.4.4.3.5 and 6.4.4.3.6 represent residuals in catch-at-age and surveys for both versions of ISVPA. Residuals in catch-at-age for the "catch-controlled" version are lower. This indicates that estimated abundances are in closer agreement with the reported catches. Residuals in catch-at-age of the "effort-controlled" version display cohort effects. This effect indicates that estimates of abundance are less based on reported catch-at-age and more based on modelled values, derived from the estimated selection pattern. Cohort effects are usually regarded as undesirable, especially when catch-at-age data are assumed to be measured without error.

In the report of ICES Working Group on methods of stock assessment (2002) obvious outliers were demonstrated in the blue whiting catch-at-age data. In cohort analyses, where calculations of abundance are dependent on the reported catches, such outliers can influence abundance estimates at other ages. In models where abundance estimation is influenced by reported catches, tracing outliers in catch residuals is hindered by adaptation of the estimates of fishing mortality to reported catches. However it does not mean that these estimates are tending towards true values. In fact the opposite is true. For abundance estimates that are based on modelled catch-at-age data (through estimated selection pattern), the cohort effects in catch residuals may outline errors in the catch-at-age data. Thus higher residuals, indicate cohorts that are less influenced by reported catch data.

If the purpose of an analysis is only to achieve better fit to reported catch-at-age data, and there are no outliers in the catch data, then a cohort effect is not desirable. However, if we are dealing with real data with noise and outliers, it is necessary to test the robustness of estimates, not merely to get a good fit to bad data.

These considerations may be supported by the fact that abundance estimates, produced by the "catch-controlled" version, which has lower residuals in catch-at-age, has lower coherence with survey data. This is demonstrated by less distinct minima of loss function profiles for corresponding surveys (compare shapes of curves 2 and 6 on figures 6.4.4.3.1 and 6.4.4.3.2).

Results of retrospective runs may be regarded as an additional support in favor of more robust effort-controlled version (figure 6.4.4.3.7): results of effort-controlled version are apparently historically more stable.

The above described factors:

- theoretically higher resistance to outliers in catch-at-age data,
- better shaped minima for signals from surveys, as well as
- more stable estimates of SSB and F in retrospective analysis
made it possible to choose the effort-controlled version of the ISVPA model - the same version which was used for blue whiting stock assessment in previous years for the final run by means of ISVPA.

Results of the ISVPA final run are presented in tables 6.4.4.3.1-6.4.4.3.3 and on figure 6.4.4.3.8. Figure 6.4.4.3.9 represents bootstrap-derived estimates of uncertainty in the results.

These results indicate a slight increase in SSB in 2003 with respect to SSB value in 2002. The results are also in line with results, obtained by this model in the previous year. The value of SSB for 2002 in 2003 stock assessment was estimated as 6.090 million tons and the estimate for 2002 in this year assessment by means of ISVPA is equal to 6.190 million tons (see figure 6.4.4.3.10). This year estimates of fishing mortality are substantially higher in comparison to last year assessment. But as it is shown in figure 6.4.4.3.4, the model in this year assessment treats the 2002 catch data as underreported. In last year assessment this year was the terminal and its catch data were regarded by the model as true. The question of whether to transfer the probable errors in the catches in terminal year into errors in stock estimates, or into errors in fishing mortality estimates (compare stable fishing mortalities, but change in trend in SSB for AMCI 2003 and 2004 assessments) - is an interesting problem.

### 6.4.4.4 Data exploration with ICA

In order to explore the data, a number of runs were made using the ICA program. Initially, the same settings were used as last year, using the newly available survey and catch at age data. The analysis was planned to investigate the effect of:

- Splitting the acoustic survey series into more recent and earlier periods
- Iterative or manual down weighting
- Length of the separable period
- Reference age for separable constraint
- Choice of reference age for separable constraint
- Down weighting of age 1 in the catch numbers
- Fixed or varying exploitation pattern
- Removal of ages from the tuning fleets

Exploratory runs were made using the same procedure as last year. The results of the later runs are presented in Table 6.4.4.4.1. In all cases, the model fit was poor, with poor definition of the minima in the SSQ surfaces.

The effect of splitting of the acoustic surveys into two series each, as has been carried out by the WG for some years, was tested by comparing run 10 and 12. Splitting the three series did not alter the residual pattern or the model fit appreciably. However, the split produced a higher estimate of SSB and a lower estimate of F.

Choice of iterative down weighting was tested by comparison of runs 11 and 12. Iterative down weighting did not improve the model fit, but did produce a higher estimate of F , in the final year, than using the same settings as in 2003. SSB estimates were not changed, however. The WG decided to continue the procedure of manual weighting, in subsequent runs.

The effect of using differing separable periods was tested by comparison of runs 13 and 14. Using a longer separable period improved the residual pattern somewhat. However in all cases a strong year effect was observed in 2002. Using a longer separable period produced slightly lower F in the final year, but also a lower SSB (Table 6.4.4.4.1).

Last year, age 6 was used as the reference age for fitting the exploitation pattern. The choice of this reference age was tested by comparing runs 13 and 25 . The change in reference age did not improve the model diagnostics, but produced a lower estimate of F , though SSB was not changed appreciably. It was considered more correct to use age 3, the first fully recruited age in the fishery. This change produced a different exploitation pattern to that of last year, and the WG considered the new pattern to be more realistic. This pattern was maintained in subsequent runs.

Varying the exploitation pattern was tested by comparing runs 12 and 15 . There was no improvement in the model fit or residual pattern, but splitting the exploitation pattern in two periods, produced higher F and lower SSB estimates.

For consistency with settings used for AMCI, age 1 fish were down weighted in the catch numbers, by $50 \%$. Comparing runs 16 b and 20 b tested this effect. The diagnostics were not improved by applying the down weighting, nor were the estimates of SSB or F altered. The effect of down weighting of age 1 by a higher value ( $90 \%$ ) was tested by comparing run 20 b and 24 . Again, there was not an appreciable change in either diagnostics or outputs.

At the meeting to evaluate assessment methods on this stock (SGAMHBW ICES 2004/ACFM24), it was recommended to remove certain ages from the tuning series. These ages were as follows:

- Norwegian acoustic survey, age 8
- Russian acoustic survey, age 7 and 8
- Norwegian Sea survey, ages 5, 6 and 7.

The effect of removing these ages was tested for two separable periods by comparing runs 22 b and 26 . There was some improvement in the diagnostics, upon removal of these ages. However, the removal of the ages produced considerably higher estimates of SSB and lower estimates of $F$ in the final year.

Finally, the effect of excluding the 2004 survey data was tested by comparing runs 19 (without 2004) and 25 (with 2004). There was no appreciable improvement in the diagnostics, but the inclusion of the 2004 survey produced a higher estimate of SSB and lower estimate of F in the final year.

The final run (run 26) used age 3 as reference age with an 8 year separable period and age 1 fish were down weighted by $50 \%$. For comparison with the inputs used elsewhere, no ages were removed from the tuning fleets. Diagnostics for this run are presented in Figure 6.4.4.4.1, 6.4.4.4.2 and 6.4.4.4.5 and outputs are given in Table 6.4.4.4.2 and Figure 6.4.4.4. Modelled catch numbers and fishing mortalities from this final ICA run are presented in Table 6.4.4.4.3 and Table 6.4.4.4.4. Confidence intervals in the estimates of F, SSB and recruitment from this run are also presented in Figure 6.4.4.4. It can be seen that uncertainty has shown a increasing trend in most recent years, especially since 1999.

| Settings used for ICA final run | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ |
| :--- | :---: | :---: |
|  |  |  |
| Number of age structured tuning series | 3, each split | 3, each split |
| Number of biomass tuning series | 0 | 0 |
| Number of years for separable constraint | 5 | 8 |
| Reference age for separable constraint | 6 | 3 |
| Constant exploitation pattern | Yes | Yes |
| S to be fixed on last age | 1.5 | 1.5 |
| Age range for mean F | $3-7$ | $3-7$ |
| Catchability model for tuning fleets | Linear | Linear |
| Age range for the analysis | $1-10$ | $1-10$ |
| Survey weights for all fleets | $100 \%$ | $100 \%$ |
| Shrinkage | No | No |
| Manual down weighting | Yes | Yes |
| Tuning series split | Yes | Yes |
| Weighting of age 1 catch numbers | $100 \%$ | $50 \%$ |

In 2004, the WG attempted to evaluate the behaviour of the ICA program to a variety of model inputs. The final run represented a slight improvement in diagnostics over pervious runs. To aid comparisons, the final run used settings were as close to AMCI, ISVPA and SMS as possible. These settings produced similar estimates of SSB and recruitment in the final year (Figure 6.4.5.3.1), though the estimate of F is somewhat lower.

### 6.4.4.5 Data exploration with SMS

SMS (Stochastic Multi Species model) is an age structured assessment model to handle biological interaction, however, it can be reduced to operate with one species only. In "single species mode" an objective functions for catch at age numbers and CPUE (or acoustic) at age time series are minimized assuming a log-normal distribution for both data sources. SMS uses maximum likelihood to weight the various data sources, and the objective function for the catch observation is then defined as presented below:.
$L_{C}=\prod_{a, y,} \frac{1}{\sigma_{\text {catch }}(a a) \sqrt{2 \pi}} \exp \left(-(\ln (C(a, y))-\ln (\hat{C}(a, y)))^{2} /\left(2 \sigma_{\text {catch }}^{2}(a a)\right)\right)$
where
C is catch at age number
$\mathrm{C}^{\wedge}$ is expected catch at age number
y is year
$a$ is age group
aa is one or more age groups

SMS is a "traditional" - forward running assessment model where the expected catch is calculated from the catch equation and F at age, which is assumed to be separable into an age selection pattern and a year effect.

The expected CPUE numbers are calculated as the product of an assumed age (or age group) dependent catchability and the mean stock number in the year. The likelihood for CPUE(survey,s,a) observations, $\mathrm{L}_{\mathrm{S}}$, is similar to $\mathrm{L}_{\mathrm{C}}$, as both are assumed lognormal distributed. The total likelihood is the product of the likelihood of the catch and the likelihood for CPUE ( $\mathrm{L}=\mathrm{L}_{\mathrm{C}} * \mathrm{~L}_{\text {CPUE }}$ ). Parameters are estimated from an minimization of $-\log (\mathrm{L})$.

The estimated model parameters include stock numbers the first year, recruitment, selection pattern and year effect for the separable catch at age model, and catchability at age for CPUE time series. Standard errors of the parameters are estimated from the Hessian matrix. SMS is implemented using the ad-model builder (Otter Research Ltd.), which is software package to create non-linear statistical models. The SMS model is still under development, but the "single species part" has extensively been tested in the last year on both simulated and real data (Vinther pers. com.).

Various configurations with respect to inclusion of data sources and options were tried with SMS, all giving an increase in both F and SSB for the last assessment year. The presented data configuration has been chosen as it is similar to the one used by the AMCI model.

## Catch data

Catch data for the period 1981-2003, age 1 to age $10+$ were used. The age selection pattern was assumed constant within two periods, 1981-1992 and 1993-2003. Selection at age was estimated by individual age group for age 1 to 7 and combined for age 8-10. Variance of catch observation was estimated separately for age 1, age 2, age 3-6 combined and age 7-10 combined.

## Tuning data

The survey data for the SMS were the same as used for the AMCI model runs. The presented AMCI runs assume that catchability is changed in the middle of the survey period. SMS assumes a constant catchability throughout a time series, but a change in catchability was obtained by splitting the three surveys into six independent surveys. The configuration with respect to time periods, catchability at age and its variance is presented below.

| Survey | Year range | Catchability at age | Variance at age of survey <br> observation |
| :--- | :--- | :--- | :--- |
| Norwegian <br> survey | acoustic | $1981-1990$ <br> $1991-2003$ | By age-group: age 2-6. <br> Combined: age 7-8 |
| Russian <br> survey acoustic | $1982-1991$ | By age: age 2, <br> Combined: 3-5 and 6-8 <br> Combined: age 7-8 |  |
| Norwegian Sea | $1992-1996$ | Combined: 3-5 and 6-8 |  |
|  | $1991-2001$ | By age-group: age 1-5. <br> Combined: age 6-7 | By age: 1-2 and 5 <br> Combined: 3-4 |

## Results

The log-likelihood contributions for the individual times series show that most information is from catch data (Table 6.4.4.5.1). The likelihood contributions are in general lower from the surveys and very poor from the Norwegian Sea survey. The same information can be found in the residuals plot (Figure 6.4.4.5.1) where the catch residuals are clearly smaller than the survey residuals. Catch residuals for the older ages shows a consistent shift for the last two years indicating that the separability assumption might be problematic. Survey residuals have a very clear year effect, often seen for acoustic time series.

The estimated selection pattern (Figure 6.4.4.5.2) indicates a shift towards younger age groups in the most recent separable period. The survey catchabilities at age (Table 6.4.4.5.2) are considerably lower in the most recent period for the older age-groups. Observed and predicted yield are fairly overlapping, however with some discrepancy for the years 2000-2002 (Figure 6.4.4.5.3).

Estimated stock numbers are presented in Table 6.4.4.5.3), fishing mortality in Table 6.4.4.5.4, and stock summary in Table 6.4.4.5.5 and Figure 6.4.5.1.1. The uncertainty of estimated stock numbers in 2003, SSB and F are presented in Figure 6.4.4.5.4. A CV at $45 \%$ for the estimated 1 -group stock number and $30 \% \mathrm{CV}$ for the 2 -group emphasize that rather little is known about the abundance of the recruiting ages to the fishery. There is a steep increase in the uncertainties of SSB in 2003 (CV=22\%) compared to SSB in 2002 (CV=14\%).

### 6.4.4.6 Comparison of assessment results and surveys

### 6.4.4.6.1 Estimates of SSB

Spawning stock biomass indices (stock estimates normalized to zero mean and unity variance) obtained from Norwegian and Russian spawning stock surveys as well as analytical assessments with AMCI, ICA, ISVPA and SMS are compared in Figure 6.4.4.6.1. All analytical assessments show fluctuating SSB indices up to mid 1990s followed by an upward trend. The annual fluctuation in the survey SSB estimates is much larger than the SSB signal suggested by
the analytical assessments up to mid 1990s. However, after mid 1990s the Norwegian acoustic survey index displays a similar increase as the analytical assessments with the exception of 2001 which is an outlier in the time series. All analytical assessments show an increase from 2002 to 2003 , whereas the survey shows a modest decrease. This discrepancy is hardly a source of concern given the strong year effects in the earlier parts of the survey time series.

### 6.4.4.6.2 Estimates of recruitment

The assessment and surveys give a consistent perception on the strength of year classes 2000 and 2001. The present assessment suggests that year class 2000 is one of the strongest year classes observed in the time series. This year class has appeared as a strong one in the majority of surveys from age 1 years onwards (Table 6.4.4.6.2.1). Year class 2001 also appears as a strong one, but most surveys suggest it is not as strong as year class 2000.

Strength of year class 2002 (as recruits in 2003) was not estimated in the final AMCI assessment, but estimation was attempted in the exploratory runs with various packages. These estimates were uncertain and mostly driven by catch at age 1 years in 2003, and suggested year class 2002 to be a strong one. Survey information does not support this result: at age 0 , it was very weak in all surveys, whereas at age 1 years it appears as an intermediate year class.

Both catch and survey information on year class 2003 are still very limited. Catches at age 0 in 2003 do not contain any clear signal. Survey information (Table 6.4.4.6.2.1) gives a mixed picture (from weak to exceptionally strong). Nevertheless, there is a slight dominance of signals suggesting this year class to be a strong one, giving reasons for cautious optimism.

### 6.4.5 Stock assessment

### 6.4.5.1 Comparison between different assessments

The results of the final assessments carried out with the AMCI, ICA, ISVPA and SMS models are compared in Figure 6.4.5.1.1. In general, the different assessments are in better agreement with each other compared to previous years.

SSB: All assessment models show good agreement in the historical estimates of SSB except for the most recent years. However, the trends remain the same. From 1998 onwards each time series indicate a considerable continued increase in SSB with a maximum in 2003. The estimates of SSB by the different assessments in 2003 range between 4 and 6 million $t$ and seems to be well above $\mathbf{B}_{\mathrm{pa}}$. The lowest SSB is estimated by AMCI and is associated with lower estimates of recruitment compared to the other models. The increase in SSB in the last 3 year is not in line with estimates of SSB estimates from the acoustic survey in spawning season, which indicates a high but constant SSB in the last 3 years (Figure 6.4.4.6.1).

Fishing mortality: The historical pattern and level of fishing mortalities are estimated to be the same by the different models. Fishing mortally increased since 1994 at or close to the highest level observed in 2003, with the exception of ICA and ISVPA which estimate a decrease in $F$ in the last year. The estimates of $F$ in 2003 by AMCI, ISVPA and SMS are well above $\mathbf{F}_{\mathrm{pa}}$ and close to $\mathbf{F}_{\text {lim }}$.

Recruitment: All models suggest an in increase in the overall level of recruitment after 1995. As well the patterns as the values of recruitment at age 1 are estimated by AMCI, ICA and SMS and are almost identical up to 1999. Thereafter the pattern remains the same but the values diverge. The recruitment estimated by ISVPA, however, differs from the other models. From 1996 onwards until 2001, ISVPA shows a continued increase in recruitment. Year class 2002 is estimated by ICA, ISVPA and SMS from the catch of this year class as 1 year old in 2003. The estimates of this year class are therefore very uncertain. AMCI has not attempted to estimate this year class and the value presented is assumed.

Catches: All models estimate catches in those periods where separabilty of the fishing pattern has been assumed. The estimates of modelled catches are compared with the reported ones in Figure 6.4.5.1.1. The figure shows that there is in general good agreement between observed and modelled catches in the historical period. However, there are differences are in the most recent period. All models estimate a higher catch in 2002 than the reported one and AMCI and ICA predict a lower catch than reported in 2003. ISVPA also show noticeable discrepancies between observed and modelled catches in some earlier years.

The overall picture, given by the different models is the same, an increase in SSB and an increase in fishing mortality in recent years, with the exception of ICA and ISVPA which suggest a decrease in F in 2003. The WG felt it difficult to make a choice between the models. There is little information supporting a specific choice between them. Nevertheless, the WG felt it necessary to make a choice as a basis for the predictions.

All models appear to be sensitive for the inclusion of a new year of data and the results differ from last year. Log catch ratio of the cohorts indicate that there may be a shift in the exploitation pattern between 2002 and 2003 (not shown).

As observed in previous years, there seems to be a conflict between the information from the catches and the surveys. The surveys contributed most to the total SSQ in the analyses. In general the minimum SSQ of the models attempted is poor defined. The diagnostics from all models show large residuals between the observed and predicted indices for all surveys. There appear to be very strong year effects in all surveys and similar in all models. The catch residuals show no trends with the exception of ISVPA where clear cohort effects exist.

It was noted that the AMCI software version, used this year was different from last year and had apparently been corrected for observed deficiencies. This resulted in a different assessment (considerably lower SSB and higher F) as last year, using last years data (Figure 6.4.5.2.7). The SMS model was used for the first time and the results of this assessment were very close to AMCI. The ISVPA model used was the same as last year, but gave quite different results. Last year a decrease in fishing mortality was estimated. This year the estimates of $F$ by ISVPA are in line with AMCI and SMS. ISVPA catch residuals seem to be very dependent on year class (either all positive or all negative) and both ISVPA and SMS do not use the 2004 data. Only AMCI and ICA use data from the 2004 acoustic survey on the spawning grounds.

Taking the considerations above into account the Working Group preferred the assessment by AMCI as a basis for the predictions.

### 6.4.5.2 Final assessment with AMCI

### 6.4.5.2.1 Configuration and diagnostics

The key settings and data for the final blue whiting assessment in 2004 are shown in the table below. The key settings of the final assessment in 2002-2003 are also shown for comparison. Some of the settings are described in more detail after the table.

| Settings/options for the AMCI run | 2002 | 2003 | 2004 |
| :---: | :---: | :---: | :---: |
| AMCI version | 2.1 | $2.2{ }^{1}$ | 2.3a |
| Age range for the analysis | 0-10+ | 1-10+ | 1-10+ |
| Last age a plus-group? | Yes | Yes | Yes |
| Age at recruitment (from Jan 1 in the year of spawning) | 0.5 | 1 | 1 |
| Recruitment in the terminal year | Fixed | Fixed | Fixed |
| Recruitment in the terminal year-1 | Estimated | Fixed | Estimated |
| Catch data |  |  |  |
| Weights for the partial objective functions for the catch fleet |  |  |  |
| Sum of squares of log catches-at-age | 1 | 1 | 1 |
| Sum of squares of log yearly yields | 1 | 1 | 1 |
| Weights of catch-at-age, age 0 and 1 years | $0.1,0.5$ | n.a., 0.5 | n.a., 0.5 |
| Gain factors in selection pattern for ages 0 and 1 years | 0.5, 0.2 | n.a., 0.2. | n.a., 0.5 |
| Gain factors in selection pattern for older ages | 0.1 | 0.1 | 0.1 |
| Selectivity for age 10 equals average of selectivity at age 8-9? | No | Yes | Yes |
| Age-structured tuning time series |  |  |  |
| Norwegian acoustic survey on the spawning grounds, ages 2-8 | 1981-2002 | 1981-2003 | 1981-2004 |
| Time series split in two periods ( $\leq 1990$ and after) | Yes | Yes | Yes |
| Flat selectivity for ages 6-8? | Yes | No | No |
| Weight in tuning for the partial objective function | 1 | 1 | 1 |
| Russian acoustic survey on the spawning grounds, ages 3-8, | 1982-1996 | 1982-1996 | 1982-1996 |
| Time series split in two periods ( $\leq 1991$ and after) | Yes | Yes | Yes |
| Flat selectivity for ages 7-8? | Yes | No | No |
| Weight in tuning for the partial objective function | 1 | 1 | 1 |
| Norwegian Sea acoustic survey, ages 1-7 | 1981-2001 | 1981-2001 | 1981-2001 |
| Time series split in two periods ( $\leq 1990$ and after) | Yes | Yes | Yes |
| Flat selectivity for ages 5-7? | Yes | No | No |
| Weight in tuning for the partial objective function | 1 | 1 | 1 |
| CPUE time series from Spanish pair trawlers, ages 1-6 ${ }^{2}$ | 1983-2001 | not used | not used |
| Flat selectivity for ages 5-6? | Yes | n.a. | n.a. |
| Weight in tuning for the partial objective function | 1 | n.a. | n.a. |
| Biomass tuning time series | 0 | 0 | 0 |

Survey data used in tuning are shown in Table 6.4.5.1.1. As in previous years, the three acoustic surveys were split into two time periods reflecting a likely change in catchability caused by a change in acoustic equipment (from Simrad EK400 to EK-500). Survey indices are treated as relative abundance indices.

Fishing mortality was modelled as separable, but with an allowance for a gradual change in the selection from year to year. The gain factor for change in selection was 0.5 for age 1 , and 0.1 for the older ages. This implies that the selection at age 1 is allowed to vary more according to the year-to-year variation in the catches than the selection at the older ages. Selection at ages 6-10 years was assumed constant (in 2003 selection was estimated for all age groups except for $10+$ years that was taken as average of two selection at two previous ages).

Recruitment as $1-y r$ old fish in 2003 was set to $20 \times 10^{9}$. This is similar to geometric mean recruitment in the period 1993-2002 and in accordance with what is known on this year class from surveys.

The key results are presented in Tables 6.4.5.2.2-6.4.5.2.4 and summarized in Figure 6.4.5.2.1. Residuals of the model fit are shown in Figure 6.4.4.2.4. Some cohort effects are visible in the catch residuals for the early cohorts. More importantly, there are blocks of positive and negative residuals for the latest years (i.e., systematic differences in modelled and observed catches), suggesting changes in selection pattern that were not tracked by the model. Year effects occur throughout the survey time series. The minimum SSQ (between the fitted model and the observed data) is not very well defined: a range of terminal F's can give SSQ that are close to the global minimum (Figure 6.4.5.2.2). Thus, the data do not allow unique characterisation of the stock in the most recent years. Selection pattern in terms of age- and year-specific F's is shown in Figure 6.4.5.2.3.

[^2]
### 6.4.5.2 $\quad$ Trends in fishing mortality, SSB and recruitment

The results of the final assessment (Table 6.4.5.2.4) indicate that fishing mortality has steadily increased since mid1990s. The exploitation pattern shows some fluctuations with the major exploitation being on adults and a sharp increase for age 1 years in 2003. SSB has been at a historically high level since 1999 and continues to display an increasing trend. Recruitment of the 2000 year class is the highest in the time series. All year classes 1995-2001 are strong, indicating a possible change in recruitment dynamics of the stock: even the weakest year class born after 1996, that of 1998, is much stronger than was typical before 1995.

A bootstrap run (Figure 6.4.5.2.4) gives an indication of the uncertainty in the assessment. Even though temporal patterns in recruitment, spawning stock biomass and fishing mortality are reproduced, uncertainty in the absolute level of these metrics during the recent years is clearly visible. In particular, the development of fishing mortality and spawning stock biomass is uncertain. The estimates of terminal F and SSB are negatively correlated (Figure 6.4.5.2.5).

### 6.4.5.2.3 Retrospective analysis and comparison with last years assessment

Retrospective analysis (Figure 6.4.5.2.6) shows that the pattern in recruitment and the generally upward trend in fishing mortality would have been rather consistently estimated in earlier assessments, but that the trend in spawning stock biomass is unstable. In comparison to the final assessment in 2003 WG , the patterns in recruitment and fishing mortality are similar but fishing mortality is at a higher level in 2004 (Figure 6.4.5.2.7). However, spawning stock biomass is estimated lower in 2004 and with a qualitatively different trend.

### 6.5 Short-term Projections

Last year the short-term prognoses by the WG were based on a deterministic prediction using MFDP. However, ACFM considered the point estimates of stock size and fishing mortality by the blue whiting assessment too uncertain as a basis for such a prognoses and based its advice on a methodology which addresses the uncertainty and the interdependence between the estimates of SSB and F.

The Working Group decided to present the forecast in a similar way as was carried out by ACFM, last year. The shortterm projections were based on a stochastic simulation using a non-parametric bootstrap of the residuals of the model fit to the catch and survey data and was carried out using STPR3 (Skagen 2002). For the population in 2004 the bootstrap run of the AMCI assessment was used. In 2004 a fishing mortality of $\mathbf{F}_{\mathrm{sq}}\left(=\mathrm{F}_{2003}\right)$ was assumed. For 2005 the prognoses was based on several assumptions of the catch level in that year. The input for the prediction is given in Table 6.5.1. The results in terms of probabilities achieving a certain fishing mortality in 2005 and surviving stock in 2006 are shown in Table 6.5 .2 with 25 and $75 \%$ confidence intervals.

The confidence intervals estimated by the model are likely to be underestimates of the true uncertainty because the recruitment estimates of the 2002 year class have been assumed to be without uncertainty.

### 6.6 Medium-term Projection

Medium-term projections were carried out with STPR version 3, i.e., the same software as used in short-term projection. The input files were generated with AMCI and were based on the final AMCI run and corresponding bootstrap replicates. The projections were based on bootstrap replicates of stock numbers in the beginning of 2004, status quo fishing mortality in 2004, historic maturity ogive and natural mortality and stock and catch weights-at-age, and selectivity estimated for 2003. Recruitment in 2004 and forward was based on Ockham's razor stock and recruitment model with variance parameter $\sigma=0.684$.

In general, medium-term projections in the most recent period are largely defined by the presently assumed stock size, recruitment and exploitation level, while the period further in the future is mostly affected by assumptions on recruitment dynamics and exploitation level. Three scenarios for both exploitation level and recruitment were considered (Table 6.6.1). Annual catches were assumed to be either 1.0, 1.5 or 2.0 million tonnes, or the highest possible catch when the target catch could not be taken. Mean recruitment level in the Ockham's razor recruitment model was chosen to be either the value obtained from the final AMCI fit (yielding mean recruitment at $13.8 \times 10^{9}$ individuals), geometric mean of recruitment estimated for the period 1993-2002 (18.7 x $10^{9}$ individuals) or arithmetic mean of recruitment estimated for the period 1993-2002 ( $23.3 \times 10^{9}$ individuals).

The projections suggest that the highest exploitation level is not sustainable for any of the considered recruitment levels, and that the lowest exploitation level is sustainable in the medium term for all of the considered recruitment levels (Table 6.6.1). Intermediate exploitation scenario was only sustainable for the highest considered recruitment level.

### 6.7 Precautionary Reference Points

The present precautionary reference points have been introduced in the advice of ACFM in 1998. The values and their technical basis are:

| Reference <br> point | $\mathbf{B}_{\text {lim }}$ | $\mathbf{B}_{\mathrm{pa}}$ | $\mathbf{F}_{\text {lim }}$ | $\mathbf{F}_{\mathrm{pa}}$ |
| :--- | :--- | :--- | :--- | :--- |
| value | 1.5 mill t | 2.25 mill. t | 0.51 | 0.32 |
| basis | $\mathbf{B}_{\text {loss }}$ | $\mathbf{B}_{\text {lim }} * 1.5$ | $\mathbf{F}_{\text {loss }}$ | $\mathbf{F}_{\text {med }}$ |

Although problems have been identified with these reference points they have remained unchanged since then. A major problem is that fishing at $\mathbf{F}_{\mathrm{pa}}$ implies a high probability of bringing the stock below $\mathbf{B}_{\mathrm{pa}}$, in other words the present combination of $\mathbf{F}_{\mathrm{pa}}$ and $\mathbf{B}_{\mathrm{pa}}$ is inconsistent.

The present reference points are based on assessments of the stock carried out at the end of the 1990s and are based on a relatively short time series. Since then major changes have been observed in the stock and in the fishery. This adds to the need to revise the reference points, because of their role as targets for rebuilding and guidelines for future exploitation.

The SGPRP revisited the $\mathbf{B}_{\text {lim }}$ reference point for blue whiting in February 2003 (ICES 2003/ACFM:15). The current $\mathbf{B}_{\text {lim }}$ a value of 1.5 million $t$ was based on an estimate of $\mathbf{B}_{\text {loss }}$ from an assessment in 1998. Since a segmented regression on the stock recruitment data was not significant, $\mathbf{B}_{\text {loss }}$ remains the obvious candidate for $\mathbf{B}_{\text {lim }}$. In the assessment carried out in $2002 \mathbf{B}_{\text {loss }}$ was estimated to be 1.2 million $t$ and SGPRP proposed this value for $\mathbf{B}_{\text {lim }}$.

Based on assessments carried out in 2004 by WGNPBW, the $\mathbf{B}_{\text {loss }}$ is estimated between 1.35 million t (AMCI), 1.7 million $t$ (ISVPA) and 1.55 million $t$ (ICA and SMS). The estimates of $\mathbf{B}_{\text {loss }}$ values in recent years vary considerable, however, around the present $\mathbf{B}_{\mathrm{lim}}$. The relatively large changes in the blue whiting assessments and the differences in stock dynamics estimated by different assessment methods make it difficult to estimate a stable value for $\mathbf{B}_{\mathrm{lim}}$. If the value of $\mathbf{B}_{\text {lim }}$ is to be revised, it is likely to be around the region of the present value of 1.5 million tonnes

It should be noted that the PA reference points presently applied in the ICES advice are based on an ICA assessment in 1998. Since then regular changes have been made in the assessment by selecting other assessment methods or different assessment configurations.

### 6.8 Management considerations

The fishery for blue whiting has expanded rapidly in recent years, while no agreement on TAC has been reached. The reported catches in 1998 to 2003 were all well above 1 million $t$ reaching a historic high in 2003 of 2.3 million $t$. The SSB has been at a fairly high level, due to continued exceptionally good recruitment in recent years. The year classes 1995-2001 are all well above average strength and there are indications that the 2003 year class also may be strong. Without the strong year classes recruited in recent years, the intensive fishery would have led to a severe depletion of the stock before now.

However, the assessments made in 2004 give a more optimistic view of the development of the stock compared to those made in 2001-2003. The main reason is that the incoming year classes are seemingly stronger than estimated previously. The SSB in 2003 is, according to all alternative assessments made, above the $\mathbf{B}_{\mathrm{pa}}$ presently used.

### 6.9 Quality of catch data and biological data

The extent of discarding in the fishery is not known. It is assumed that any discarding that occurs in the industrial fishery is confined to net bursting through overloading or through the catch exceeding the RSW tank capacity. On
freezer vessels landing a catch for human consumption the role of grading in handling the catch is not known. If grading occurs in this sector then landing figures may not reflect the whole catch.

The type and frequency of by-catch in the fishery has not been sufficiently explored. Each fishery could be expected to have a different by-catch profile. The submission of by catch data for each of the blue whiting fisheries would enable the WG to report on the likely impacts of blue whiting fisheries on other stocks and fisheries.

The text table below shows the number of samples and total landings by the three fisheries, Directed, Mixed, and Southern, and by quarter.

| Quarter |  | Directed | Mixed | Southern | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Number of samples | 111 | 8 | 115 | 234 |
|  | Landings (t) | 541258 | 12082 | 3905 | 557245 |
| 2 | Number of samples | 346 | 39 | 130 | 515 |
|  | Landings (t) | 868935 | 43235 | 4195 | 916365 |
| 3 | Number of samples | 385 | 21 | 142 | 548 |
|  | Landings (t) | 528913 | 57628 | 3982 | 590523 |
| 4 | Number of samples | 133 | 20 | 146 | 299 |
|  | Landings (t) | 237941 | 40752 | 4394 | 283087 |
| Total | Number of samples | 975 | 88 | 533 | 1596 |
|  | Landings (t) | 2177047 | 153697 | 16476 | 2347220 |

In total 1596 samples were collected from the fisheries, 188,770 fish measured and 26,207 fish aged. Sampled fish were not evenly distributed throughout the fisheries. The most intensive sampling took place in the south. Here one sample was taken for every 31 tonnes, followed by the mixed fishery with one sample for every 1747 tonnes, and lastly the directed fishery where there was one sample for every 2233 tonne caught. In this context it should be noted that implementation of the EU Collection of Fisheries Data, Fisheries Regulation 1639/2001, requires a minimum of one sample to be taken for every 1000 t landed. Detailed information on the number of samples, number of fish measured, and number of fish aged by country and quarter is given in Tables 6.9.1 and 6.9.2. As can be seen, no sampling was carried out by France, Scotland, and Sweden.

The WG requires the samples to estimate catch in numbers and mean length and mean weight. Therefore, the WG urges all countries that exploit this stock to develop appropriate sampling schemes.

## $6.10 \quad$ Recommendations

- It is recommended to initiate research on the recruitment mechanism of blue whiting and to seek an answer to the question why the productivity of blue whiting in the northeast Atlantic, through higher recruitment, has increased during the last 10 years.
- Comparison of age distributions from surveys and landings originating from different research institutes indicate that there are differences in the interpretation of the structure in the otoliths resulting in different age estimates. The WG recommends that an otolith exchange programme will be carried out in 2004 which makes it possible to investigate the magnitude of the problem.
- It is recommended that existing information on discards and by-catch in the fisheries is made available to the WG.
- The WG considers that the basic knowledge of the blue whiting could be strengthened, and it is therefore recommended that more internationally coordinated basic research should be carried out, including surveying the total geographical distribution and studying the structure (age and size distribution) of blue whiting in various sub-regions throughout the year.

It would be helpful to the working group's deliberations if countries participating in the fishery could enumerate the number, capacity and effort of vessels prosecuting the fishery.
Table 6.2.1 Landings (tonnes) of BLUE WHITING from the directed fisheries (Sub-areas I and II, Division Va, XIVa and XIVb) 1987-2003, as estimated by the Working

| Country | 1987 | 1988 | $1989{ }^{3)}$ | 1990 | 1991 | 1992 | 1993 | 1994 ${ }^{2 /}$ | $1995^{3)}$ | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | - | - | . | - | - | - | - | . | - | - | - | - | 15 | 7,721 | 5,723 | 13,608 | 38,226 |
| Estonia | - | - | - | - | - | - | - | - | - | 377 | 161 | 904 | - | - | - | - |  |
| Faroes | 9,290 | - | 1,047 | - | - | - | - | - | - | 345 | - | 44,594 | 11,507 | 17,980 | 64,496 | 82,977 | 115,755 |
| Germany | 1,010 | 3 | 1,341 | - | - | - | - | 2 | 3 | 32 | - | 78 | - | - | 3117 | 1,072 | 813 |
| Greenland | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Iceland | - | - | 4,977 | - | - | - | - | - | 369 | 302 | 10,464 | 68,681 ${ }^{\text {4 }}$ | 96,295 | 155,024 | 245,814 | 195,483 | 312,334 |
| Latvia | - | - | - | - | - | - | - | 422 | - | - | - | - | - | - | - | - |  |
| Netherlands | - | - | - | - | - | - | - | - | 72 | 25 | - | 63 | 435 | - | 5180 | 906 | 592 |
| Norway ${ }^{\text {s }}$ | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 64,581 | 100,922 | 215,075 |
| Norway ${ }^{\text {( }}$ | - | - | - | 566 | 100 | 912 | 240 | - | - | 58 | 1,386 | 12,132 | 5,455 | - | 28,812 | - | - |
| Poland | 56 | 10 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| Sweden | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 850 | 57,206 |
| USSR/ Russia ${ }^{11}$ | 112,686 | 55,816 | 35,250 | 1,540 | 78,603 | 61,400 | 43,000 | 22,250 | 23,289 | 22,308 | 50,559 | 51,042 | 65,932 | 103,941 | 173,860 | 145,649 | 191,507 |
| Total | 123,042 | 55,829 | 42,615 | 2,106 | 78,703 | 62,312 | 43,240 | 22,674 | 23,733 | 23,447 | 62,570 | 177,494 | 179,639 | 284,666 | 591,583 | 541,467 | 931,508 |

${ }^{1}$ ) From 1992 only Russia
2) Includes Vb for Russia.
${ }^{3}$ ) Icelandic mixed fishery in Va.
${ }^{4}$ ) include mixed in Va and directed in Vb .
Directed fishery
By-catches of blue whiting in other fisheries.
Table 6.2.2 Landings (tonnes) of BLUE WHITING from directed fisheries (Division Vb,VIa,b, VIIb,c. VIIg-k and Sub-area XII) 1987-2003 as estimated by the Working Group.

| Country | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | $1998{ }^{1)}$ | 1999 | 2000 | 2001 | 2002 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 2,655 | 797 | 25 | - | - | 3,167 | - | 770 | - | 269 | - | 5051 | 19,625 | 11,856 | 18,110 | 2,141 | 17,813 |
| Estonia | - | - | - | - | - | 6,156 | 1,033 | 4,342 | 7754 | 10,605 | 5,517 | 5,416 | - | - | - | - | - |
| Faroes | 70,625 | 79,339 | 70,711 | 43,405 | 10,208 | 12,731 | 14,984 | 22,548 | 26,009 | 18,258 | 22,480 | 26,328 | 93,234 | 129,969 | 188,464 | 115,127 | 208,427 |
| France | - | - | 2,190 | - | - | - | 1,195 | - | 720 | 6,442 | 12,446 | 7,984 | 6,662 | 13,481 | 13,480 | 14,688 | 13,365 |
| Germany | 3,850 | 5,263 | 4,073 | 1,699 | 349 | 1,307 | 91 | - | 6,310 | 6,844 | 4,724 | 17,891 | 3,170 | 12,655 | 15,862 | 15,378 | 21,866 |
| Iceland | - | - | - | - | - | - | - | - | - | - | - | - | 64,135 | 105,833 | 119,287 | 91,853 | 189,159 |
| Ireland | 3,706 | 4,646 | 2,014 | - | - | 781 | - | 3 | 222 | 1,709 | 25,785 | 45635 | 35,240 | 25,200 | 29,854 | 17,723 | 22,484 |
| Japan | - | - | - | - | - | 918 | 1,742 | 2,574 | - | - | - | - | - | - | - | - | - |
| Latvia | - | - | - | - | - | 10,742 | 10,626 | 2,160 | - | - | - | - | - | - | - | - | - |
| Lithauen | - | - | - | - | - | - | 2,046 | - | - | - | - | - | - | - | - | - | - |
| Netherlands ${ }^{2}$ ) | 5,627 | 800 | 2,078 | 7,280 | 17,359 | 11,034 | 18,436 | 21,076 | 26,703 | 17,644 | 23,676 | 27,884 | 35,408 | 46,128 | 68,415 | 33,365 | 45,239 |
| Norway | 191,012 | 208,416 | 258,386 | 281,036 | 114,866 | 148,733 | 198,916 | 226,235 | 261,272 | 337,434 | 318,531 | 519,622 | 475,004 | 460,274 | 399,932 | 385,495 | 502,320 |
| UK (Scotland) | 3,315 | 5,071 | 8,020 | 6,006 | 3,541 | 6,849 | 2,032 | 4,465 | 10,583 | 14,325 | 33,398 | 92,383 | 98,853 | 42,478 | 50,147 | 26,403 | 27,136 |
| Sweden | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 10 | - |
| USSR/ Russia ${ }^{3}$ ) | 165,497 | 121,705 | 127,682 | 124,069 | 72,623 | 115,600 | 96,000 | 94,531 | 83,931 | 64,547 | 68,097 | 79,000 | 112,247 | 141,257 | 141,549 | 144,419 | 163,812 |
| Total | 446,287 | 426,037 | 475,179 | 463,495 | 218,946 | 318,018 | 347,101 | 378,704 | 423,504 | 478,077 | 514,654 | 827,194 | 943,578 | 989,131 | 1,045,100 | 846,602 | 1,211,621 |

[^3]${ }^{3}$ ) From 1992 only Russia
Table 6.2.3 Landings (tonnes) of BLUE WHITING from directed fisheries and by-catches caught in other fisheries in Divisions IIIa, IVa 1987-2003, as estimated by the WG.

| Country | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | $1993{ }^{3)}$ | 1994 | 1995 | 1996 | 1997 | $1998{ }^{\text {2) }}$ | 1999 | 2000 | 2001 | 2002 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark ${ }^{4)}$ |  |  | 3,632 | 10,972 | 5,961 | 4,438 | 25,003 | 5,108 | 4,848 | 29,137 | 9,552 | 40,143 | 36,492 | 30,360 | 21,995 |  |  |
| Denmark ${ }^{5}$ | 28,541 | 18,144 | 22,973 | 16,080 | 9,577 | 26,751 | 16,050 | 14,578 | 7,591 | 22,695 | 16,718 | 16,329 | 8,521 | 7,749 | 7,505 | 35,530 | 26,896 |
| Faroes ${ }^{4)}{ }^{\text {9 }}$ |  |  |  |  |  |  |  |  |  |  |  | - | - | - | 60 |  |  |
| Faroes 5) ${ }^{\text {a }}$ | 7,051 | 492 | 3,325 | 5,281 | 355 | 705 | 1,522 | 1,794 | - | 6,068 | 6,066 | 296 | 265 | 42 | 6,741 | 7,317 | 5,712 |
| Germany ${ }^{1)}$ | 115 | 280 | 3 | - | - | 25 | 9 | - | - | - | - |  |  | - | 81 | - | 36 |
| Ireland | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 4 |  |
| Netherlands | - | - | - | 20 | - | 2 | 46 | - | - | - | 793 |  |  | - | - | 50 | 0 |
| Norway ${ }^{4}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 21,804 |  |  |
| Norway ${ }^{5}$ | 24,969 | 24,898 | 42,956 | 29,336 | 22,644 | 31,977 | 12,333 | 3,408 | 78,565 | 57,458 | 27,394 | 28,814 | 48,338 | 73,006 | 58,182 | 85,062 | 117,145 |
| Russia | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 69 | - | - |
| Sweden | 2,013 | 1,229 | 3,062 | 1,503 | 1,000 | 2,058 | 2,867 | 3,675 | 13,000 | 4,000 | 4,568 | 9,299 | 12,993 | 3,319 | 2,086 | 17,689 | 8,326 |
| UK | - | 100 | 7 | - | 335 | 18 | 252 | - | - | 1 | - | - | - | - | - | - | 65 |
| Total | 62,689 | 45,143 | 75,958 | 63,192 | 39,872 | 65,974 | 58,082 | 28,563 | 104,004 | 119,359 | 65,091 | 94,881 | 106,609 | 114,476 | 118,523 | 145,652 | 158,180 |
| ${ }^{1}$ ) Including directed fishery also in Division IVa. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{2}$ ) Including mixed industrial fishery in the Norwegian Sea |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{3}$ ) Imprecise estimates for Sweden: reported catch of 34265 t in 1993 is replaced by the mean of 1992 and 1994, i.e. 2,867t, and used in the assessment. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{4)}$ Directed fishery |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{5}$ ) By-catches of blue whiting in other fisheries. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ${ }^{6}$ ) For the periode 1987-2000 landings figures also include landings from mixed fisheries in Division Vb. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 6.2.4 Landings (tonnes) of BLUE WHITING from the Southern areas (Sub-areas VIII and IX and Divisions VIIg-k and VIId,e) 1987-2003, as estimated by the Working Group.

| Country | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Germany | - | - | - | - | - | - | - | - | - | - | - | - |  |  |  | $600{ }^{2)}$ | $88^{2)}$ |
| Ireland | - | - | - | - | - | - | - | - | - | - | - | - |  |  |  | $98^{2)}$ | $96^{2}$ |
| Netherlands | - | - | - | 450 | 10 | - | - | - | - | - | - | $10^{1)}$ |  |  |  | $3208{ }^{2)}$ | 2471, ${ }^{2}$ |
| Norway | 4 | - | - | - | - | - | - | - | - | - | - |  |  | - | - | - |  |
| Portugal | 9,148 | 5,979 | 3,557 | 2,864 | 2,813 | 4,928 | 1,236 | 1,350 | 2,285 | 3,561 | 2,439 | 1,900 | 2,625 | 2,032 | 1,746 | 1,659 | 2,651 |
| Spain | 23,644 | 24,847 | 30,108 | 29,490 | 29,180 | 23,794 | 31,020 | 28,118 | 25,379 | 21,538 | 27,683 | 27,490 | 23,777 | 22,622 | 23,218 | 17,506 | 13,825 |
| UK | 23 | 12 | 29 | 13 | - | - | - | 5 | - | - | - | - | - | - | - | - | 181 |
| France | - | - | 1 | - | - | - | - | - | - | - | - | - | - | - | - | - | 784 |
| Total | 32,819 | 30,838 | 33,695 | 32,817 | 32,003 | 28,722 | 32,256 | 29,473 | 27,664 | 25,099 | 30,122 | 29,400 | 26,402 | 24,654 | 24,964 | 23,071 | 20,097 |

Landings reported as Directed fisheries and included in the Catch-at-Age calculations of that fisheries

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| I |  |  |  |  |  |  |  |  | 46 |  |  |  |  | 46 |
| IIa | 38,226 | 105,563 |  | 813 | 240,273 |  | 215,075 |  | 174,425 |  |  | 57,206 | 592 | 832,173 |
| IIb |  |  |  |  |  |  |  |  | 16,973 |  |  |  |  | 16,973 |
| IIIa | 7,316 | 1,243 |  |  |  |  |  |  |  |  |  | 7,603 |  | 16,162 |
| IVa | 18,886 | 4,137 |  | 36 |  |  | 117,145 |  |  | 65 |  | 723 | 0 | 140,992 |
| IVb | 694 | 332 |  |  |  |  |  |  |  |  |  |  | 0 | 1,026 |
| IXa |  |  |  |  |  |  |  | 2,651 |  |  |  |  |  | 2,651 |
| Va |  | 10,192 |  |  | 72,061 |  |  |  |  |  |  |  |  | 82,253 |
| Vb | 9,439 | 138,435 | 1,572 | 8,104 | 180,568 |  | 16,483 |  | 104,630 | 30 |  |  | 9,008 | 468,269 |
| VIa | 5,044 | 30,297 |  | 9,573 |  | 1 | 72,332 |  |  | 3,253 |  |  | 5,537 | 126,037 |
| VIb | 985 | 10,361 | 3,256 | 907 | 7,060 | 5,619 | 221,028 |  | 55,716 | 20,441 |  |  | 12,542 | 337,915 |
| VIIb |  |  |  |  |  | 25 |  |  |  | 251 |  |  |  | 276 |
| VIIbc |  |  | 3,592 | 3,282 |  |  | 121,216 |  |  |  |  |  |  | 128,090 |
| VIIc |  | 11,094 |  |  |  | 8,586 |  |  |  | 3,160 |  |  | 18,010 | 40,850 |
| VIIg |  |  |  |  |  | 4 |  |  |  |  |  |  |  | 4 |
| VIIgk |  |  | 4,945 |  |  |  | 71,262 |  |  |  |  |  |  | 76,207 |
| VIIgk+XII |  | 14,357 |  |  | 1,531 |  |  |  | 3,466 |  |  |  |  | 19,354 |
| VIIh |  |  |  |  |  |  |  |  |  |  |  |  | 142 | 142 |
| VIIIabd |  |  | 784 |  |  |  |  |  |  |  |  |  | 596 | 1,380 |
| VIIIc+IXa |  |  |  |  |  |  |  |  |  |  | 13,825 |  |  | 13,825 |
| VIIj |  |  |  | 88 |  | 96 |  |  |  | 181 |  |  | 1,876 | 2,241 |
| VIIk |  | 3,884 |  |  |  | 6,740 |  |  |  |  |  |  | 0 | 10,624 |
| XII | 2,345 |  |  |  |  | 1,509 |  |  |  |  |  |  |  | 3,854 |
| XIVb |  |  |  |  |  |  |  |  | 63 |  |  |  |  | 63 |
| Grand Total | 82,935 | 329,895 | 14,149 | 22,803 | 501,493 | 22,580 | 834,540 | 2,651 | 355,319 | 27,382 | 13,825 | 65,532 | 48,303 | 2,321,406 |

Table 6.2.6 Landings (tonnes) of BLUE WHITING from the main fisheries, 1987-2003, as estimated by the Working Group.

| Area | Norwegian Sea fishery <br> (Sub-areas $1+2$ and <br> Divisions Va, XIVa-b) | Fishery in the spawning area (Divisions Vb, VIa, VIb and VIIb-c) | Directed- and mixed fisheries (Divisions III: and IV ) | Total northern areas | Total southern areas (Subareas VIII and IX and Divisions VIId, e, $\mathrm{g}-\mathrm{k})$ | Grand total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 123,042 | 446,287 | 62,689 | 632,018 | 32,819 | 664,837 |
| 1988 | 55,829 | 426,037 | 45,143 | 527,009 | 30,838 | 557,847 |
| 1989 | 42,615 | 475,179 | 75,958 | 593,752 | 33,695 | 627,447 |
| 1990 | 2,106 | 463,495 | 63,192 | 528,793 | 32,817 | 561,610 |
| 1991 | 78,703 | 218,946 | 39,872 | 337,521 | 32,003 | 369,524 |
| 1992 | 62,312 | 318,081 | 65,974 | 446,367 | 28,722 | 475,089 |
| 1993 | 43,240 | 347,101 | 58,082 | 448,423 | 32,256 | 480,679 |
| 1994 | 22,674 | 378,704 | 28,563 | 429,941 | 29,473 | 459,414 |
| 1995 | 23,733 | 423,504 | 104,004 | 551,241 | 27,664 | 578,905 |
| 1996 | 23,447 | 478,077 | 119,359 | 620,883 | 25,099 | 645,982 |
| 1997 | 62,570 | 514,654 | 65,091 | 642,315 | 30,122 | 672,437 |
| 1998 | 177,494 | 827,194 | 94,881 | 1,099,569 | 29,400 | 1,128,969 |
| 1999 | 179,639 | 943,578 | 106,609 | 1,229,826 | 26,402 | 1,256,228 |
| 2000 | 284,666 | 989,131 | 114,477 | 1,388,274 | 24,654 | 1,412,928 |
| 2001 | 591,583 | 1,045,100 | 118,523 | 1,755,206 | 24,964 | 1,780,170 |
| 2002 | 541,467 | 846,602 | 145,652 | 1,533,721 | 23,071 | 1,556,792 |
| 2003 | 931,508 | 1,211,621 | 158,180 | 2,301,309 | 20,097 | 2,321,406 |

Table 6.2.7 Total landings of blue whiting by quarter and area for 2003 in tonnes. Landing figures provided by Working Group members.

| Area | Quarter 1 | Quarter 2 | Quarter 3 | Quarter 4 | Grand Total |
| :--- | ---: | ---: | ---: | ---: | ---: |
| I | 46 |  |  |  | 46 |
| IIa | 6,494 | 356,694 | 389,168 | 79,816 | 832,173 |
| IIb | 6,508 | 10,465 |  |  | 16,973 |
| IIIa | 24 | 2,002 | 12,230 | 1,906 | 16,162 |
| IVa | 13,778 | 42,769 | 55,446 | 28,999 | 140,992 |
| IVb | 25 | 0 | 1,001 |  | 1,026 |
| IXa | 586 | 734 | 603 | 728 | 2,651 |
| Va | 1,326 | 9,116 | 51,460 | 20,351 | 82,253 |
| Vb | 19,153 | 242,563 | 68,757 | 137,795 | 468,269 |
| VIa | 3,107 | 117,839 | 813 | 4,278 | 126,037 |
| VIb | 220,002 | 116,505 | 1,408 |  | 337,915 |
| VIIb | 255 | 21 |  |  | 276 |
| VIIbc | 128,090 |  |  |  | 128,090 |
| VIIc | 39,031 | 1,819 |  |  | 40,850 |
| VIIg |  | 4 |  |  | 4 |
| VIIgk | 76,207 |  |  |  | 76,207 |
| VIIgk+XII | 18,979 | 375 |  |  | 19,354 |
| VIIh | 142 |  |  |  |  |
| VIIIabd | 221 | 6 | 1,153 |  | 142 |
| VIIIc+IXa | 3,319 | 3,461 | 3,379 | 3,666 | 1,380 |
| VIIj | 2,056 | 113 | - |  | 72 |

Table 6.2.8 Scottish Blue Whiting Fleet 2003

|  | Quarter 1 | Quarter 2 |
| :--- | ---: | ---: |
| No of Vessels | 12 |  |
| Fishing Gear | Industrial Trawl | Industrial Trawl |
| Length Range Vessels (m) | $51-74$ | $51-74$ |
| Mean Length Vessels (m) | 60 | 59 |
| Vessel Type | Pelagic Trawler | Pelagic Trawler |
| Power range(kw) | $1700-6210$ | $1700-4330$ |
| Mean Power (KW) | 2353 | 2017 |
| Average crew size | 10 | 10 |
| Catch Market | Fish Meal | Fish Meal |
| No of Vessels in Pelagic Fleet | 33 |  |
| Approx Annual Value (£000's) | 2000 |  |

Table 6.3.1.1 Blue whiting. Landing in numbers ('000) by length group (cm) and quarters for the Nothern area in 2003.

| Length | Q1 | Q2 | Q3 | Q4 | All year |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 5 |  |  |  |  |  |
| 6 |  |  |  |  |  |
| 7 |  |  |  |  |  |
| 8 |  |  |  |  |  |
| 9 |  |  |  |  |  |
| 10 |  |  |  |  |  |
| 11 |  |  |  |  |  |
| 12 |  | 1 |  |  | 1 |
| 13 |  | 2 | 996 | 10 | 1009 |
| 14 | 2 | 33 | 1093 | 20 | 1149 |
| 15 | 53 | 443 | 161 | 819 | 1477 |
| 16 | 551 | 795 | 43 | 996 | 2385 |
| 17 | 658 | 75382 | 3106 | 2402 | 81549 |
| 18 | 636 | 296292 | 8602 | 1148 | 306678 |
| 19 | 1122 | 585536 | 18315 | 2261 | 607234 |
| 20 | 9307 | 422709 | 67563 | 26551 | 526130 |
| 21 | 30376 | 278637 | 128766 | 54021 | 491800 |
| 22 | 84753 | 401772 | 215736 | 108496 | 810756 |
| 23 | 203226 | 587015 | 412634 | 139752 | 1342626 |
| 24 | 509897 | 798318 | 575120 | 188154 | 2071488 |
| 25 | 652430 | 697924 | 581374 | 195229 | 2126957 |
| 26 | 642613 | 745563 | 491912 | 158284 | 2038372 |
| 27 | 465444 | 468176 | 301613 | 108754 | 1343989 |
| 28 | 392580 | 427543 | 210312 | 88149 | 1118584 |
| 29 | 235319 | 313269 | 159687 | 44790 | 753065 |
| 30 | 207968 | 219153 | 91570 | 31199 | 549891 |
| 31 | 153237 | 108838 | 56758 | 23809 | 342642 |
| 32 | 79028 | 61254 | 33470 | 13519 | 187272 |
| 33 | 65828 | 26774 | 10559 | 7200 | 110362 |
| 34 | 29047 | 18271 | 9306 | 7047 | 63671 |
| 35 | 13122 | 14216 | 3832 | 2063 | 33233 |
| 36 | 11072 | 10790 | 4 | 595 | 22461 |
| 37 | 8017 | 10058 | 961 | 64 | 19100 |
| 38 | 4763 | 5429 | 5 | 41 | 10237 |
| 39 | 2 | 2604 | 1 | 22 | 2629 |
| 40 | 4579 | 3481 | 1 | 14 | 8076 |
| 41 |  | 296 |  | 22 | 318 |
| 42 |  | 666 |  | 1 | 667 |
| 43 |  | 665 |  |  | 665 |
| 44 |  | 9 |  |  | 9 |
| 45 |  |  |  |  |  |
| 46 |  |  |  |  |  |
| 47 |  |  |  |  |  |
| 48 |  |  |  |  |  |
| 49 |  |  |  |  |  |
| 50 |  |  |  |  |  |
| TOTAL numbers | 3805632 | 6581915 | 3383499 | 1205435 | 14976480 |


| Official Catch (t) | 563721 | 867968 | 522534 | 237800 | 2192024 |
| :--- | :--- | :--- | :--- | :--- | :--- |

Table 6.3.1.2 Blue whiting. Landing in numbers ('000) by length group ( cm ) and quarters for the North Sea and Skagerrak in 2003.

| Length | Q1 | Q2 | Q3 | Q4 | All year |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 5 |  |  |  |  |  |
| 6 |  |  |  |  |  |
| 7 |  |  |  |  |  |
| 8 |  |  |  |  |  |
| 9 |  |  |  |  |  |
| 10 |  |  |  |  |  |
| 11 |  |  |  |  |  |
| 12 | 203 |  | 505 | 408 | 1117 |
| 13 | 203 | 663 | 16865 | 5895 | 23626 |
| 14 |  | 884 | 68220 | 33412 | 102515 |
| 15 | 407 | 2651 | 47563 | 30131 | 80753 |
| 16 | 6509 | 9087 | 14024 | 21788 | 51407 |
| 17 | 10836 | 20973 | 4548 | 3673 | 40029 |
| 18 | 5862 | 49633 | 20844 | 4406 | 80745 |
| 19 | 6047 | 93089 | 78135 | 15402 | 192672 |
| 20 | 7360 | 122964 | 164478 | 47150 | 341951 |
| 21 | 14442 | 85317 | 187222 | 102863 | 389844 |
| 22 | 27090 | 54189 | 114893 | 64146 | 260319 |
| 23 | 19749 | 61897 | 69859 | 44150 | 195655 |
| 24 | 13147 | 56575 | 53373 | 35768 | 158863 |
| 25 | 8950 | 36265 | 57542 | 39706 | 142463 |
| 26 | 9116 | 21884 | 24508 | 25219 | 80727 |
| 27 | 5751 | 13911 | 15727 | 18160 | 53549 |
| 28 | 5233 | 6997 | 8653 | 11079 | 31963 |
| 29 | 3181 | 6058 | 7579 | 6529 | 23347 |
| 30 | 4198 | 4889 | 6632 | 5366 | 21084 |
| 31 | 1017 | 2366 | 2526 | 2040 | 7950 |
| 32 | 666 | 442 | 1516 | 1836 | 4460 |
| 33 | 462 | 1261 | 1011 | 1020 | 3754 |
| 34 |  | 221 | 505 | 816 | 1542 |
| 35 |  | 442 |  | 612 | 1054 |
| 36 |  |  |  |  |  |
| 37 |  |  |  | 204 | 204 |
| 38 |  |  |  |  |  |
| 39 |  |  |  |  |  |
| 40 |  |  |  |  |  |
| 41 |  |  |  |  |  |
| 42 |  |  |  |  |  |
| 43 |  |  |  |  |  |
| 44 |  |  |  |  |  |
| 45 |  |  |  |  |  |
| 46 |  |  |  |  |  |
| 47 |  |  |  |  |  |
| 48 |  |  |  |  |  |
| 49 |  |  |  |  |  |
| 50 |  |  |  |  |  |
| TOTAL numbers | 150428 | 652657 | 966729 | 521781 | 2291595 |
|  |  |  |  |  |  |
| Official Catch (t) | 12097 | 44231 | 64006 | 35495 | 155829 |

Table 6.3.1.3 Blue whiting. Landing in numbers ('000) by length group (cm) and quarters for the Southern area in 2003.

| Length | Q1 | Q2 | Q3 | Q4 | All year |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 5 |  |  |  |  |  |
| 6 |  |  |  |  |  |
| 7 |  |  |  |  |  |
| 8 |  |  |  |  |  |
| 9 |  |  |  |  |  |
| 10 |  |  |  |  |  |
| 11 |  |  |  | 4 | 4 |
| 12 |  |  |  |  |  |
| 13 |  |  |  |  |  |
| 14 |  | 13 | 22 | 2 | 37 |
| 15 | 527 | 15 | 54 | 114 | 711 |
| 16 | 3500 | 445 | 323 | 1647 | 5915 |
| 17 | 8965 | 2830 | 1102 | 2791 | 15688 |
| 18 | 6874 | 8461 | 7051 | 3583 | 25969 |
| 19 | 2898 | 8605 | 17616 | 11303 | 40422 |
| 20 | 2635 | 7230 | 12082 | 7680 | 29626 |
| 21 | 6566 | 10665 | 8684 | 7419 | 33334 |
| 22 | 7395 | 10946 | 7749 | 9961 | 36051 |
| 23 | 8137 | 8060 | 5289 | 7820 | 29306 |
| 24 | 6373 | 5738 | 3368 | 4622 | 20100 |
| 25 | 2993 | 2739 | 1956 | 3743 | 11430 |
| 26 | 2179 | 1331 | 1339 | 1944 | 6793 |
| 27 | 1275 | 1013 | 572 | 1186 | 4046 |
| 28 | 973 | 501 | 449 | 586 | 2509 |
| 29 | 616 | 211 | 156 | 471 | 1454 |
| 30 | 251 | 86 | 102 | 268 | 707 |
| 31 | 96 | 16 | 32 | 120 | 264 |
| 32 | 155 | 13 | 7 | 86 | 262 |
| 33 | 32 | 11 | 5 | 9 | 57 |
| 34 | 12 | 13 | 6 | 1 | 32 |
| 35 | 15 | 5 | 1 | 4 | 26 |
| 36 | 6 | 6 | 3 | 1 | 15 |
| 37 | 1 | 4 | 2 | 1 | 8 |
| 38 | 8 | 4 | 1 |  | 13 |
| 39 | 1 | 2 | 1 |  | 4 |
| 40 |  | 3 | 1 |  | 5 |
| 41 |  |  |  |  |  |
| 42 |  |  |  |  |  |
| 43 |  |  |  |  |  |
| 44 |  |  |  |  |  |
| 45 |  |  |  |  |  |
| 46 |  |  |  |  |  |
| 47 |  |  |  |  |  |
| 48 |  |  |  |  |  |
| 49 |  |  |  |  |  |
| 50 |  |  |  |  |  |
| TOTAL numbers | 62483 | 68965 | 67972 | 65366 | 264785 |
|  |  |  |  |  |  |
| Official Catch (t) | 3905 | 4195 | 3982 | 4394 | 16477 |

Table 6.3.2.1 BLUE WHITING. Catch in number (millions) by age group in the directed fisheries (Sub-areas I and II, Divisions Va, and XIVa+b, Vb, VIa+b, VIIbc and VIIg-k) in 1992-2003.

|  |  | 2003 |
| :---: | :---: | :---: |
|  |  | 58 |
|  |  | 2464 |
|  |  | 3626 |
|  |  | 7964 |
|  |  | 4726 |
|  |  | 2006 |
|  |  | 1090 |
|  |  | 398 |
|  |  | 119 |
|  |  | 18 |
|  |  | 27 |
|  |  | 22,495 |
|  |  | $2,177,047$ |

fisheries (Divisions IIIa and IV) for 1992-2003.

|  |  | 2003 |
| :---: | :---: | :---: |
|  |  | 222 |
|  |  | 1191 |
|  |  | 369 |
|  |  | 368 |
|  |  | 73 |
|  |  | 18 |
|  |  | 23 |
|  |  | 1 |
|  |  | 1 |
|  |  | 1 |
|  |  | 1 |
|  |  | 2,269 |



Table 6.3.2.4. Blue Whiting: Catch in numbers (thousands) of the total stock in 1981-2003

| Age |  | $\mathbf{1 9 8 1}$ | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{0}$ |  |  |  |  |  |  | $\mathbf{1 9 8 8}$ |
|  | $\mathbf{1}$ | 258,000 | 148,000 | $2,283,000$ | $2,291,000$ | $1,305,000$ | 650,000 | 838,000 |
|  | $\mathbf{2}$ | 348,000 | 274,000 | 567,000 | $2,331,000$ | $2,044,000$ | 816,000 | 578,000 |
|  | $\mathbf{3}$ | 681,000 | 326,000 | 270,000 | 455,000 | $1,933,000$ | $1,862,000$ | 728,000 |
|  | $\mathbf{4}$ | 334,000 | 548,000 | 286,000 | 260,000 | 303,000 | $1,717,000$ | $1,897,000$ |
|  | $\mathbf{5}$ | 548,000 | 264,000 | 299,000 | 285,000 | 188,000 | 393,000 | 726,000 |
|  | $\mathbf{6}$ | 559,000 | 276,000 | 304,000 | 445,000 | 321,000 | 187,000 | 137,000 |
|  | $\mathbf{7}$ | 466,000 | 266,000 | 287,000 | 262,000 | 257,000 | 201,000 | 105,000 |
|  | $\mathbf{8}$ | 634,000 | 272,000 | 286,000 | 193,000 | 174,000 | 198,000 | 123,000 |
|  | $\mathbf{9}$ | 578,000 | 284,000 | 225,000 | 154,000 | 93,000 | 174,000 | 103,000 |
|  | $\mathbf{1 0 +}$ | $1,460,000$ | 673,000 | 334,000 | 255,000 | 259,000 | 398,000 | 195,000 |


| Age |  | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{0}$ |  |  |  |  |  |  |  |  |
|  | $\mathbf{1}$ | 865,000 | $1,611,000$ | 266,686 | 407,730 | 263,184 | 306,951 | 296,100 | $1,893,453$ |
|  | $\mathbf{2}$ | 718,000 | 703,000 | $1,024,468$ | 653,838 | 305,180 | 107,935 | 353,949 | 534,221 |
|  | $\mathbf{3}$ | $1,340,000$ | 672,000 | 513,959 | $1,641,714$ | 621,085 | 367,962 | 421,560 | 632,361 |
|  | $\mathbf{4}$ | 791,000 | 753,000 | 301,627 | 569,094 | $1,571,236$ | 389,264 | 465,358 | 537,280 |
|  | $\mathbf{5}$ | 837,000 | 520,000 | 363,204 | 217,386 | 411,367 | $1,221,919$ | 615,994 | 323,324 |
|  | $\mathbf{6}$ | 708,000 | 577,000 | 258,038 | 154,044 | 191,241 | 281,120 | 800,201 | 497,458 |
|  | $\mathbf{7}$ | 139,000 | 299,000 | 159,153 | 109,580 | 107,005 | 174,256 | 253,818 | 663,133 |
|  | $\mathbf{8}$ | 50,000 | 78,000 | 49,431 | 79,663 | 64,769 | 90,429 | 159,797 | 232,420 |
|  | $\mathbf{9}$ | 25,000 | 27,000 | 5,060 | 31,987 | 38,118 | 79,014 | 59,670 | 98,415 |
|  | $\mathbf{1 0 +}$ | 38,000 | 95,000 | 9,570 | 11,706 | 17,476 | 30,614 | 41,811 | 82,521 |


| Age |  | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{0}$ |  |  |  |  |  |  |  |
|  | $\mathbf{1}$ | $2,131,494$ | $1,656,926$ | 788,200 | $1,814,851$ | $4,363,690$ | $1,821,053$ | $3,742,841$ |
|  | $\mathbf{2}$ | $1,519,327$ | $4,181,175$ | $1,549,100$ | $1,192,657$ | $4,486,315$ | $3,232,244$ | $4,073,497$ |
|  | $\mathbf{3}$ | 904,074 | $3,541,231$ | $5,820,800$ | $3,465,739$ | $2,962,163$ | $3,291,844$ | $8,378,955$ |
|  | $\mathbf{4}$ | 577,676 | $1,044,897$ | $3,460,600$ | $5,014,862$ | $3,806,520$ | $2,242,722$ | $4,824,590$ |
|  | $\mathbf{5}$ | 295,671 | 383,658 | 412,800 | $1,550,063$ | $2,592,933$ | $1,824,047$ | $2,035,096$ |
|  | $\mathbf{6}$ | 251,642 | 322,777 | 207,200 | 513,663 | 585,666 | $1,647,122$ | $1,117,179$ |
|  | $\mathbf{7}$ | 282,056 | 303,058 | 151,200 | 213,057 | 170,020 | 344,403 | 400,022 |
|  | $\mathbf{8}$ | 406,910 | 264,105 | 153,100 | 151,429 | 97,032 | 168,848 | 121,280 |
|  | $\mathbf{9}$ | 104,320 | 212,452 | 68,800 | 58,277 | 76,624 | 102,576 | 19,701 |
| $\mathbf{1 0 +}$ | 169,235 | 85,513 | 140,500 | 139,791 | 66,410 | 142,743 | 27,493 |  |

Table 6.3.3.1. Blue Whiting: Mean weights-at-age in the total catch and stock in 1981-2003

| Age |  | $\mathbf{1 9 8 1}$ | $\mathbf{1 9 8 2}$ | $\mathbf{1 9 8 3}$ | $\mathbf{1 9 8 4}$ | $\mathbf{1 9 8 5}$ | $\mathbf{1 9 8 6}$ | $\mathbf{1 9 8 7}$ | $\mathbf{1 9 8 8}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{0}$ | 0.038 | 0.018 | 0.020 | 0.026 | 0.016 | 0.030 | 0.023 | 0.031 |
|  | $\mathbf{1}$ | 0.052 | 0.045 | 0.046 | 0.035 | 0.038 | 0.040 | 0.048 | 0.053 |
|  | $\mathbf{2}$ | 0.065 | 0.072 | 0.074 | 0.078 | 0.074 | 0.073 | 0.086 | 0.076 |
|  | $\mathbf{3}$ | 0.103 | 0.111 | 0.118 | 0.089 | 0.097 | 0.108 | 0.106 | 0.097 |
|  | $\mathbf{4}$ | 0.125 | 0.143 | 0.140 | 0.132 | 0.114 | 0.130 | 0.124 | 0.128 |
|  | $\mathbf{5}$ | 0.141 | 0.156 | 0.153 | 0.153 | 0.157 | 0.165 | 0.147 | 0.142 |
|  | $\mathbf{6}$ | 0.155 | 0.177 | 0.176 | 0.161 | 0.177 | 0.199 | 0.177 | 0.157 |
|  | $\mathbf{7}$ | 0.170 | 0.195 | 0.195 | 0.175 | 0.199 | 0.209 | 0.208 | 0.179 |
|  | $\mathbf{8}$ | 0.178 | 0.200 | 0.200 | 0.189 | 0.208 | 0.243 | 0.221 | 0.199 |
|  | $\mathbf{9}$ | 0.187 | 0.204 | 0.204 | 0.186 | 0.218 | 0.246 | 0.222 | 0.222 |
|  | $\mathbf{1 0 +}$ | 0.213 | 0.231 | 0.228 | 0.206 | 0.237 | 0.257 | 0.254 | 0.260 |


| Age |  | $\mathbf{1 9 8 9}$ | $\mathbf{1 9 9 0}$ | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{0}$ | 0.014 | 0.034 | 0.036 | 0.024 | 0.028 | 0.033 | 0.022 |
|  | $\mathbf{1}$ | 0.059 | 0.045 | 0.055 | 0.057 | 0.066 | 0.061 | 0.064 |
|  | $\mathbf{2}$ | 0.079 | 0.070 | 0.091 | 0.083 | 0.082 | 0.087 | 0.091 |
|  | $\mathbf{3}$ | 0.103 | 0.106 | 0.107 | 0.119 | 0.109 | 0.108 | 0.118 |
|  | $\mathbf{4}$ | 0.126 | 0.123 | 0.136 | 0.140 | 0.137 | 0.137 | 0.143 |
|  | $\mathbf{5}$ | 0.148 | 0.147 | 0.174 | 0.167 | 0.163 | 0.164 | 0.154 |
|  | $\mathbf{6}$ | 0.158 | 0.168 | 0.190 | 0.193 | 0.177 | 0.189 | 0.167 |
|  | $\mathbf{7}$ | 0.171 | 0.175 | 0.206 | 0.226 | 0.200 | 0.207 | 0.116 |
|  | $\mathbf{8}$ | 0.203 | 0.214 | 0.230 | 0.235 | 0.217 | 0.217 | 0.206 |
|  | $\mathbf{9}$ | 0.224 | 0.217 | 0.232 | 0.284 | 0.225 | 0.247 | 0.236 |
|  | $\mathbf{1 0 +}$ | 0.253 | 0.256 | 0.266 | 0.294 | 0.281 | 0.254 | 0.256 |


| Age |  | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{0}$ | 0.031 | 0.033 | 0.035 | 0.031 | 0.038 | 0.021 |
|  | $\mathbf{1}$ | 0.047 | 0.048 | 0.063 | 0.057 | 0.050 | 0.054 |
|  | $\mathbf{2}$ | 0.072 | 0.072 | 0.078 | 0.075 | 0.078 | 0.074 |
|  | $\mathbf{3}$ | 0.102 | 0.094 | 0.088 | 0.086 | 0.094 | 0.093 |
|  | $\mathbf{4}$ | 0.121 | 0.125 | 0.109 | 0.104 | 0.108 | 0.115 |
|  | $\mathbf{5}$ | 0.140 | 0.149 | 0.142 | 0.133 | 0.129 | 0.132 |
|  | $\mathbf{6}$ | 0.166 | 0.178 | 0.170 | 0.156 | 0.163 | 0.155 |
|  | $\mathbf{7}$ | 0.177 | 0.183 | 0.199 | 0.179 | 0.186 | 0.131 |
|  | $\mathbf{8}$ | 0.183 | 0.188 | 0.193 | 0.187 | 0.193 | 0.143 |
|  | $\mathbf{9}$ | 0.203 | 0.221 | 0.192 | 0.232 | 0.231 | 0.224 |
|  | $\mathbf{1 0 +}$ | 0.232 | 0.248 | 0.245 | 0.241 | 0.243 | 0.262 |

Table 6.3.4.1. Blue Whiting: natural mortality and proportion of maturation at age. Natural mortality is assumed to be the same in all years. The values for the maturity-ogive were estimated by the 1994 WG (ICES 1995/Assess:7).

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Proportion of mature | 0.00 | 0.11 | 0.40 | 0.82 | 0.86 | 0.91 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 |
| Natural mortality | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |

Table 6.4.1.1 Age stratified acoustic survey estimates of blue whiting in the spawning area west of the British Isles by Norwegian reserach vessels, March/ April 1981-2004. Number in millions.

| Numbers | age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | total |
| 1981 |  | 2372 | 7583 | 3253 | 3647 | 4611 | 4638 | 3654 | 2591 | 1785 | 34134 |
| 1982 |  | no survey |  |  |  |  |  |  |  |  |  |
| 1983 |  | 297 | 2108 | 2723 | 6511 | 3735 | 3650 | 3153 | 2279 | 1182 | 25638 |
| 1984 |  | 15767 | 1721 | 1616 | 1719 | 1858 | 1128 | 567 | 440 | 348 | 25164 |
| 1985 |  | no survey |  |  |  |  |  |  |  |  |  |
| 1986 |  | 1003 | 5829 | 4122 | 624 | 228 | 203 | 250 | 137 | 170 | 12566 |
| 1987 |  | 4960 | 8417 | 22589 | 4735 | 282 | 417 | 385 | 159 | 27 | 41971 |
| 1988 |  | 9712 | 9090 | 12367 | 20392 | 7355 | 723 | 599 | 326 | 398 | 60962 |
| 1989 |  | 6787 | 22270 | 9973 | 10504 | 7803 | 933 | 293 | 177 | 46 | 58786 |
| 1990 |  | 14169 | 12670 | 11228 | 5587 | 6556 | 3273 | 516 | 183 | 108 | 54290 |
| 1991 |  | 11147 | 6340 | 8497 | 7407 | 4558 | 2019 | 545 | 96 | 16 | 40625 |
| 1992 |  | 1232 | 26123 | 4719 | 1574 | 1386 | 810 | 616 | 257 | 19 | 36736 |
| 1993 |  | 4489 | 3321 | 26771 | 2643 | 1270 | 557 | 426 | 108 | 22 | 39607 |
| 1994 |  | 1603 | 2950 | 4476 | 11354 | 1742 | 1687 | 908 | 770 | 207 | 25697 |
| 1995 |  | 8538 | 9874 | 7906 | 6861 | 9467 | 1795 | 1083 | 482 | 149 | 46155 |
| 1996 |  | 8781 | 7433 | 8371 | 2399 | 4455 | 4111 | 1202 | 459 | 162 | 37373 |
| 1997 |  | no survey |  |  |  |  |  |  |  |  |  |
| 1998 |  | 18218 | 34991 | 4697 | 1674 | 279 | 407 | 381 | 351 | 86 | 61084 |
| 1999 |  | 19034 | 60309 | 26103 | 1481 | 316 | 72 | 153 | 141 | 0 | 107609 |
| 2000 | 12783 | 8613 | 31011 | 41382 | 6843 | 898 | 427 | 228 | 139 | 115 | 102439 |
| 2001 | 16479 | 44162 | 12843 | 13805 | 8292 | 718 | 175 | 51 | 0 | 0 | 96525 |
| 2002 | 20455 | 71996 | 54740 | 12757 | 5266 | 8404 | 1450 | 305 | 15 | 176 | 175564 |
| 2003 | 26520 | 23992 | 70303 | 28756 | 5735 | 2430 | 1708 | 260 | 229 |  | 133184 |
| 2004 | 3860 | 18569 | 40669 | 50137 | 15649 | 4454 | 2218 | 1313 | 426 |  | 137295 |

$\square$ used in the assessment

Table 6.4.1.2. Age stratified acoustic survey estimates of blue whiting in the spawning area by Russian vessels Number in millions, biomass in thousand t , mean length in cm , mean weight in grams

| Numbers | age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | total |
| 1981 |  |  |  |  |  |  |  |  |  |  |  |
| 1982 |  |  | 540 | 2750 | 1340 | 1380 | 1570 | 2350 | 1730 | 1290 | 12950 |
| 1983 |  |  | 2330 | 2930 | 9390 | 3880 | 1970 | 1370 | 780 | 660 | 23310 |
| 1984 |  |  | 2900 | 800 | 1100 | 4200 | 2200 | 1200 | 1700 | 1200 | 15300 |
| 1985 |  |  | 13220 | 930 | 580 | 1780 | 860 | 610 | 580 | 540 | 19100 |
| 1986 |  |  | 18750 | 23180 | 2540 | 610 | 620 | 750 | 640 | 710 | 47800 |
| 1987 |  |  | 4480 | 19170 | 5860 | 1070 | 500 | 810 | 860 | 670 | 33420 |
| 1988 |  |  | 3710 | 4550 | 8610 | 4130 | 1270 | 480 | 250 | 260 | 23260 |
| 1989 |  |  | 11910 | 7120 | 6670 | 6970 | 4580 | 2750 | 1880 | 810 | 42690 |
| 1990 |  |  | 9740 | 12140 | 5740 | 2580 | 1470 | 220 | 80 | 10 | 31980 |
| 1991 |  |  | 10300 | 5350 | 5130 | 2630 | 1770 | 870 | 300 | 220 | 26570 |
| 1992 |  |  | 20010 | 6700 | 1350 | 440 | 390 | 170 | 0 | 0 | 29060 |
| 1993 |  |  | 4728 | 12337 | 5304 | 2249 | 1316 | 621 | 386 | 150 | 27091 |
| 1994 |  |  | no survey |  |  |  |  |  |  |  |  |
| 1995 |  |  | 12657 | 10028 | 8942 | 2651 | 1093 | 408 | 131 | 14 | 35924 |
| 1996 |  |  | 15285 | 10629 | 4897 | 6940 | 1482 | 653 | 85 | 0 | 39971 |
| 1997 |  |  | no survey |  |  |  |  |  |  |  |  |
| 1998 |  |  | no survey |  |  |  |  |  |  |  |  |
| 1999 |  |  | no survey |  |  |  |  |  |  |  |  |
| 2000 |  |  | no survey |  |  |  |  |  |  |  |  |
| 2001 |  |  | no comparab | urvey |  |  |  |  |  |  |  |
| 2002 |  |  | no comparab | urvey |  |  |  |  |  |  |  |
| 2003 |  |  | no comparab | urvey |  |  |  |  |  |  |  |
| 2004 |  |  | no compara | urvey |  |  |  |  |  |  |  |

$\square$ used in the assessment

Table 6.4.1.3. Age stratified acoustic survey estimates of blue whiting in the Norwegian Sea in July Number in millions, biomass in thousand t , mean length in cm , mean weight in grams

| Numbers | age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | total |
| 1981 |  | 182 | 728 | 4542 | 3874 | 2678 | 2834 | 2964 | 2756 | 2054 | 22612 |
| 1982 |  | 184 | 460 | 1242 | 4715 | 3611 | 3128 | 2323 | 1679 | 874 | 18216 |
| 1983 |  | 22356 | 396 | 468 | 756 | 1404 | 576 | 468 | 432 | 324 | 27180 |
| 1984 |  | 30380 | 13916 | 833 | 392 | 539 | 539 | 343 | 49 | 49 | 47040 |
| 1985 |  | 5969 | 23876 | 12502 | 658 | 423 | 188 | 235 | 141 | 376 | 44368 |
| 1986 |  | 2324 | 2380 | 7224 | 6944 | 1876 | 952 | 336 | 308 | 140 | 22484 |
| 1987 |  | 8204 | 4032 | 5180 | 5572 | 1204 | 224 | 168 | 56 | 84 | 24724 |
| 1988 |  | 4992 | 2880 | 2640 | 3480 | 912 | 120 | 96 | 24 | 48 | 15192 |
| 1989 |  | 1172 | 1125 | 812 | 379 | 410 | 212 | 22 | 32 |  | 4164 |
| 1990 |  | no survey |  |  |  |  |  |  |  |  |  |
| 1991 |  | no survey |  |  |  |  |  |  |  |  |  |
| 1992 |  | 792 | 1134 | 6939 | 766 | 247 | 172 | 90 | 11 | 18 | 10169 |
| 1993 |  | 830 | 125 | 1070 | 6392 | 1222 | 489 | 248 | 58 | 88 | 10522 |
| 1994 |  | no survey |  |  |  |  |  |  |  |  |  |
| 1995 |  | 6974 | 2811 | 1999 | 1209 | 1622 | 775 | 173 | 61 |  | 15624 |
| 1996 |  | 23464 | 1057 | 899 | 649 | 436 | 505 | 755 | 69 | 41 | 27875 |
| 1997 |  | 30227 | 25638 | 1524 | 779 | 300 | 407 | 260 | 137 | 123 | 59395 |
| 1998 |  | 24244 | 47815 | 16282 | 556 | 212 | 100 | 64 | 10 | 255 | 89538 |
| 1999 |  | 14367 | 9750 | 23701 | 9754 | 1733 | 466 | 79 | 48 | 91 | 59989 |
| 2000 |  | 25813 | 3298 | 2721 | 3078 | 23 | 46 | 6 |  |  | 34985 |
| 2001 |  | 61470 | 22051 | 7883 | 3225 | 1824 | 156 | 12 | 0 | 68 | 96689 |
| 2002 |  | no survey |  |  |  |  |  |  |  |  | 0 |
| 2003 |  | no survey |  |  |  |  |  |  |  |  | 0 |

[^4]Table 6.4.1.1.1 Age and length distribution of blue whiting in the international survey by R/Vs "Celtic Explorer", "Fridtjof Nansen", "Johan Hjort" and "Tridens", west of the British Isles, March-April 2004.


Table 6.4.1.1.2 Age and length distribution of blue whiting in the survey by R/V "Johan Hjort" west of the British Isles, March-April 2004.


Table 6.4.1.1.3. Blue whiting biomass estimates (million tonnes) in the spawning area.

| Year | Russia |  | Norway |  | International |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | total | spawning | total | spawning | total | spawning |
| 1983 | 3.6 | 3.6 | 4.7 | 4.4 |  |  |
| 1984 | 3.4 | 2.7 | 2.8 | 2.1 |  |  |
| 1985 | 2.8 | 2.7 |  |  |  |  |
| 1986 | 6.4 | 5.6 | 2.6 | 2.0 |  |  |
| 1987 | 5.4 | 5.1 | 4.3 | 4.1 |  |  |
| 1988 | 3.7 | 3.1 | 7.1 | 6.8 |  |  |
| 1989 | 6.3 | 5.7 | 7.0 | 6.1 |  |  |
| 1990 | 5.4 | 5.1 | 6.3 | 5.7 |  |  |
| 1991 | 4.6 | 4.2 | 5.1 | 4.8 |  |  |
| 1992 | 3.6 | 3.3 | 4.3 | 4.2 |  |  |
| 1993 | 3.8 | 3.7 | 5.2 | 5.0 |  |  |
| 1994 |  |  | 4.1 | 4.1 |  |  |
| 1995 | 6.8 | 6.0 | 6.7 | 6.1 |  |  |
| 1996 | 7.1 | 5.8 | 5.1 | 4.5 |  |  |
| 1997 |  |  |  |  |  |  |
| 1998 |  |  | 5.5 | 4.7 |  |  |
| 1999 |  |  | 8.9 | 8.5 |  |  |
| 2000 |  |  | 8.3 | 7.8 |  |  |
| 2001 |  |  | 6.7 | 5.6 |  |  |
| 2002 | 5.2 |  | 12.2 | 10.9 |  |  |
| 2003 | 18.5 | 16.6 | 11.4 | 10.4 |  |  |
| 2004 | 4.3 | 4.1 | 11.4 | 10.9 | 11.1 | 10.6 |
| Mean | 5.7 | 5.2 | 6.5 | 5.9 | 11.1 | 10.6 |

Table 6.4.1.2.1. Blue Whiting. Age stratified acoustic survey estimates in the PGSPFN/PGNAPES coordinated May/June survey Norwegian Sea (not used in the assessment Estimate is for the area north of $63^{\circ} \mathrm{N}$ and east of $8^{\circ} \mathrm{W}$ only.


Table 6.4.1.2.2. Blue whiting. Age stratified acoustic survey estimates in the Icelandic EEZ in July (not used in the assessment)

| Year | Age (yrs) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9+ | total |
| Numbers (millions) |  |  |  |  |  |  |  |  |  |  |  |
| 1999 | 14869 | 2100 | 1357 | 1772 | 5790 | 1344 | 316 | 50 | 15 | 42 | 27655 |
| 2000 | 10683 | 8594 | 934 | 523 | 1218 | 468 | 106 | 25 | 1 | 1 | 22553 |
| 2001 | 27305 | 4090 | 5215 | 1657 | 1614 | 398 | 132 | 37 | 6 | 2 | 40456 |
| 2002 | 3815 | 10785 | 3107 | 1436 | 1724 | 1430 | 727 | 178 | 47 | 5 | 23254 |
| 2003 | 5011 | 9363 | 6054 | 7430 | 3888 | 1350 | 852 | 581 | 91 | 33 | 34653 |
| 2003* | 5011 | 9158 | 4899 | 4645 | 1918 | 646 | 218 | 227 | 91 | 6 | 26819 |


|  |  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |  | 9+ | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| biomass (tonnes*10^-3) |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1999 | 265 | 163 | 127 | 201 | 764 | 212 | 55 | 13 |  | 4 | 14 | 1818 |
|  | 2000 | 186 | 624 | 85 | 63 | 167 | 78 | 22 | 5 |  |  |  | 1230 |
|  | 2001 | 661 | 295 | 568 | 211 | 231 | 66 | 22 | 8 |  | 1 | 0 | 2063 |
|  | 2002 | 77 | 746 | 297 | 160 | 217 | 203 | 114 | 31 |  | 13 | 1 | 1859 |
|  | 2003 | 68 | 555 | 600 | 853 | 503 | 200 | 147 | 102 |  | 19 | 9 | 3055 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | $9+$ | total |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| mean length (cm) |  |  |  |  |  |  |  |  |  |  |  |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | $9+$ | total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mean weight (gr) |  |  |  |  |  |  |  |  |  |  |  |
| 1999 | 18 | 77 | 94 | 113 | 132 | 158 | 174 | 254 | 248 | 307 | 66 |
| 2000 | 16 | 71 | 90 | 115 | 136 | 161 | 208 | 186 | 241 |  | 53 |
| 2001 | 24 | 72 | 109 | 127 | 143 | 165 | 190 | 218 | 194 | 234 | 51 |
| 2002 | 20 | 69 | 96 | 111 | 126 | 142 | 157 | 173 | 267 | 205 | 80 |
| 2003 | 14 | 59 | 99 | 115 | 129 | 148 | 172 | 175 | 212 | 257 | 88 |

Table 6.4.2.1 Stratified mean catch ( Kg /haul and Number/haul) and standard error of BLUE WHITING in bottom trawl surveys in Spanish waters (Divisions VIIIc and IXa north). All surveys in September-October.

| Kg/haul | 30-100 m |  | 101-200 m |  | 201-500 m |  | TOTAL 30-500 m |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | Mean | SD | Mean | SD | Mean | SD |
| 1985 | 9.50 | 5.87 | 119.75 | 45.99 | 68.18 | 13.79 | 92.83 | 28.24 |
| 1986 | 9.74 | 7.13 | 45.41 | 12.37 | 29.54 | 8.70 | 36.93 | 7.95 |
| 1987 | - |  | - |  | - |  | - |  |
| 1988 | 2.90 | 2.59 | 154.12 | 38.69 | 183.07 | 141.94 | 143.30 | 45.84 |
| 1989 | 14.17 | 12.03 | 76.92 | 17.08 | 18.79 | 6.23 | 59.00 | 11.68 |
| 1990 | 6.25 | 3.29 | 52.54 | 9.00 | 18.80 | 4.99 | 43.60 | 6.60 |
| 1991 | 64.59 | 34.65 | 126.41 | 26.06 | 46.07 | 18.99 | 97.10 | 17.16 |
| 1992 | 6.37 | 2.59 | 44.12 | 6.64 | 29.50 | 6.16 | 34.60 | 4.23 |
| 1993 | 1.06 | 0.63 | 14.07 | 3.73 | 51.08 | 22.02 | 22.59 | 6.44 |
| 1994 | 8.04 | 5.28 | 37.18 | 8.45 | 25.42 | 5.27 | 29.70 | 5.19 |
| 1995 | 19.97 | 13.87 | 36.43 | 4.82 | 15.97 | 4.10 | 28.52 | 3.66 |
| 1996 | 7.27 | 3.95 | 49.23 | 7.19 | 92.54 | 17.76 | 54.52 | 6.36 |
| 1997 | 7.60 | 4.44 | 44.21 | 10.61 | 60.18 | 17.54 | 44.01 | 8.00 |
| 1998 | 5.29 | 1.92 | 41.09 | 7.64 | 73.80 | 24.06 | 44.48 | 7.82 |
| 1999 | 31.41 | 7.28 | 108.46 | 17.24 | 150.24 | 39.53 | 108.12 | 14.62 |
| 2000 | 39.52 | 9.73 | 88.89 | 14.32 | 62.23 | 27.65 | 74.42 | 11.25 |


|  | $70-120 \mathrm{~m}$ |  | $121-200 \mathrm{~m}$ |  | $201-500 \mathrm{~m}$ |  | TOTAL 70-500 m |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1997 | 17.87 | 7.35 | 44.68 | 10.52 | 57.14 | 16.60 | 42.62 |
| 1998 | 14.13 | 4.17 | 42.78 | 8.13 | 78.88 | 22.01 | 47.14 | 7.58 |
| 1999 | 92.66 | 14.60 | 11.76 | 19.87 | 169.21 | 50.26 | 124.27 | 17.83 |
| 2000 | 62.39 | 12.00 | 91.99 | 14.75 | 58.72 | 24.94 | 76.19 | 10.61 |
| 2001 | 8.35 | 3.31 | 50.18 | 10.09 | 52.41 | 16.71 | 42.02 | 7.02 |
| 2002 | 31.40 | 5.02 | 69.00 | 13.41 | 36.75 | 12.07 | 51.80 | 7.64 |
| 2003 | 42.52 | 12.22 | 71.40 | 11.01 | 46.43 | 11.42 | 58.13 | 6.92 |


| Number/haul | 30-100 m |  | 101-200 m |  | 201-500 m |  | TOTAL 30-500 m |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | Mean | SD | Mean | SD | Mean | SD |
| 1985 | 267 | 181.71 | 3669 | 1578.86 | 1377 | 262.98 | 2644 | 963.20 |
| 1986 | 368 | 237.56 | 2486 | 1006.67 | 752 | 238.87 | 1763 | 616.40 |
| 1987 | - |  | - |  | - |  | - |  |
| 1988 | 83 | 71.74 | 6112 | 1847.36 | 7276 | 6339.88 | 5694 | 2086.00 |
| 1989 | 629 | 537.29 | 3197 | 876.75 | 566 | 213.11 | 2412 | 599.00 |
| 1990 | 220 | 115.48 | 2219 | 426.46 | 578 | 185.43 | 1722 | 276.00 |
| 1991 | 2922 | 1645.73 | 5563 | 1184.69 | 1789 | 847.33 | 4214 | 780.88 |
| 1992 | 124 | 50.81 | 1412 | 233.99 | 845 | 199.12 | 1069 | 146.87 |
| 1993 | 14 | 8.61 | 257 | 69.61 | 894 | 427.77 | 401 | 124.53 |
| 1994 | 346 | 234.12 | 2002 | 456.50 | 997 | 245.91 | 1487 | 689.00 |
| 1995 | 1291 | 864.97 | 2004 | 341.48 | 485 | 137.81 | 1493 | 240.37 |
| 1996 | 147 | 82.71 | 1167 | 167.20 | 2097 | 385.23 | 1263 | 142.30 |
| 1997 | 224 | 121.69 | 1425 | 359.12 | 1254 | 330.37 | 1228 | 234.50 |
| 1998 | 123 | 44.12 | 1442 | 334.24 | 1823 | 592.92 | 1347 | 251.37 |
| 1999 | 795 | 218.58 | 3996 | 697.66 | 5279 | 1521.62 | 3861 | 576.10 |
| 2000 | 1574 | 360.78 | 3701 | 568.17 | 2036 | 857.01 | 2940 | 406.62 |


|  | $70-120 \mathrm{~m}$ |  | $121-200 \mathrm{~m}$ |  | $201-500 \mathrm{~m}$ |  | TOTAL $70-500 \mathrm{~m}$ |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 552 | 235.60 | 1443 | 361.89 | 1183 | 323.14 | 1180 | 209.94 |
| 1998 | 351 | 105.96 | 1463 | 320.26 | 2012 | 590.04 | 1387 | 234.82 |
| 1999 | 2502 | 427.23 | 4358 | 847.87 | 6119 | 2026.39 | 4474 | 727.32 |
| 2000 | 2267 | 414.97 | 3930 | 604.11 | 2009 | 859.71 | 3027 | 400.87 |
| 2001 | 171 | 77.34 | 1310 | 263.84 | 1232 | 381.49 | 1048 | 172.74 |
| 2002 | 771 | 90.34 | 2526 | 499.30 | 1075 | 331.09 | 1739 | 268.70 |
| 2003 | 1320 | 384.25 | 2791 | 554.16 | 1513 | 454.02 | 2114 | 317.68 |

Table 6.4.2.2 BLUE WHITING. Stratified mean catch ( $\mathrm{Kg} / \mathrm{haul}$ ) and standard error of in bottom trawl surveys in Portuguese waters (Division IXa).

| Year | Month | 20-100 m |  | 100-200 m |  | 200-500 m |  | 500-750 m |  | TOTAL: 20-500 m |  | TOTAL: 20-750 m |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | y | sy | y | sy | y | sy | y | sy | y | sy | y | sy |
| 1979 | June | 0 |  | 33 |  | 86 | 35 | - |  | 31 | 12 | - |  |
|  | Oct./Nov. | 5 | 5 | 17 | 8 | 103 | 48 | - | - | 28 | 9 | - | - |
| 1980 | March | 0 | 0 | 178 | 173 | 5 | 1 | - | - | 72 | 69 | - |  |
|  | May/June | 1 |  | 4 |  | 45 | 18 | - | - | 11 | 4 | - |  |
|  | October | 4 | 3 | 10 | 4 | 587 | 306 | - | - | 117 | 58 | - | - |
| 1981 | March | 0 | 0 | 24 | 17 | 186 | 113 | - |  | 42 | 22 |  |  |
|  | June | 0 | 0 | 4 | 2 | 178 | 25 | - | - | 34 | 4 | - | - |
| 1982 | April/May | 0 | 0 | 3 | 3 | 136 | 39 | - |  | 26 | 7 |  |  |
|  | September | 1 | 1 | 85 | 42 | 271 | 123 | - | - | 86 | 29 | - | - |
| 1983 | March | 1 | 1 | 14 | 10 | 259 | 96 | - | - | 54 | 18 | - |  |
|  | June | 0 | 0 | 23 | 8 | 177 | 47 | - | - | 42 | 9 | - | - |
| 1985 | June | 0 | 0 | 194 | 146 | 405 | 162 | - | - | 159 | 68 | - |  |
|  | October | 4 | 3 | 133 | 84 | 341 | 39 | - | - | 120 | 35 | - | - |
| 1986 | June | 4 | 1 | 59 | 19 | 196 | 31 | - | - | 65 | 10 | - |  |
|  | October | 2 | 1 | 357 | 144 | 650 | 111 | - | - | 276 | 63 | - | - |
| 1987 | October | 3 | 0 | 297 | 64 | 747 | 229 | - | - | 263 | 50 | - | - |
| 1988 | October | 4 | 2 | 165 | 47 | 457 | 106 | - | - | 155 | 28 | - |  |
| 1989 | July | 0 | 0 | 42 | 21 | 323 | 143 | 79 | 36 | - | - | 78 | 24 |
|  | October | 7 | 4 | 70 | 26 | 306 | 84 | 24 | 2 | - | - | 79 | 16 |
| 1990 | July | 2 | 2 | 153 | 103 | 242 | 42 | 50 | 5 | - | - | 96 | 35 |
|  | October | 11 | 5 | 90 | 28 | 762 | 234 | 42 | 10 | - | - | 153 | 35 |
| 1991 | July | 1 | 1 | 140 | 40 | 268 | 38 | 64 | 18 | - | - | 98 | 15 |
|  | October | 8 | 5 | 83 | 18 | 259 | 53 | 121 | 27 | - | - | 91 | 11 |
| 1992 | February | 7 | 7 | 43 | 35 | 249 | 21 | 73 | 3 |  |  | 68 | 12 |
|  | July | 1 | 1 | 29 | 18 | 216 | 43 | 27 | 5 | - | - | 47 | 9 |
|  | October | 1 | 1 | 22 | 7 | 208 | 44 | 80 | 3 | - | - | 54 | 7 |
| 1993 | February | 0 | 0 | 19 | 14 | 105 | 31 | 36 | 0 | - | - | 42 | 10 |
|  | July | 0 |  | 3 | 3 | 151 | 28 | 55 | 5 | - | - | 34 | 4 |
|  | November | 0 | 0 | 90 | 0 | 189 | 43 | 6 | 1 | - | - | 86 | 9 |
| 1994 | October | 0 | 0 | 374 | 30 | 283 | 32 | 49 | 7 | - | - | 174 | 11 |
| 1995 | July | 0 | 0 | 18 | 14 | 130 | 20 | 52 | 3 | - | - | 35 | 5 |
|  | October | 18 | 15 | 103 | 21 | 328 | 91 | 31 | 12 | - | - | 94 | 16 |
| 1996 | October | 25 | 24 | 12 | 2 | 36 | 6 | 25 | 7 |  |  | 22 | 8 |
| 1997 | June | 0 | 0 | 3 | 3 | 116 | 42 | 45 | 12 | - | - | 27 | 7 |
|  | October | 2 | 1 | 54 | 20 | 77 | 13 | 7 | 2 | - | - | 32 | 8 |
| 1998 | July | 0 |  | 8 | 5 | 105 | 17 | 38 | 3 | - | - | 25 | 3 |
|  | October | 1 | 1 | 384 | 87 | 427 | 101 | 20 | 2 | - | - | 212 | 36 |
| 1999 | July | 1 |  | 60 |  | 66 |  | 25 |  | - | - | 37 | n/a |
|  | October | 0 |  | 70 |  | 78 |  | 18 |  | - | - | 41 | n/a |
| 2000 | July | 23 | 13 | 109 | 34 | 116 | 10 | 63 | 6 | - |  | 75 | 13 |
|  | October | 11 | 4 | 155 | 53 | 196 | 22 | 54 | 4 | - | - | 100 | 19 |
| 2001 | July | 18 | 7 | 238 | 37 | 305 | 37 | 57 | 14 | - | - | 152 | 23 |
|  | October | 106 | 6 | 474 | 224 | 294 | 66 | 0 | 0 | - | - | 295 | 97 |
| 2002 | October | 19 | 12 | 176 | 81 | 180 | 24 | 0 | 0 | - | - | 116 | 34 |
| 2003 | October | 24 | 10 | 114 | 14 | 119 | 30 | 34 | 6 | - |  | 76 | 8 |

Table 6.4.2.3. Catch in number by age of blue whiting from the summer (August/September) bottom trawl surveys on the Faroe plateau 1996-2003. Icelandic age readings from June-August were used to split the numbers by age.

|  |  |  |  |  | Age |  |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | Total |  |  |
| 1996 | 12513 | 18586 | 4576 | 5392 | 6754 | 2755 | 1610 | 768 | 352 | 337 | 121 | 2 | 34 | 53802 |  |  |
| 1997 | 4139 | 20745 | 13710 | 8345 | 5748 | 2488 | 1376 | 619 | 242 | 179 | 95 | 2 | 14 | 57701 |  |  |
| 1998 | 2359 | 21202 | 28278 | 19217 | 12289 | 4143 | 2330 | 1057 | 358 | 301 | 126 | 4 | 27 | 91690 |  |  |
| 1999 | 7322 | 4189 | 4468 | 12725 | 19609 | 6041 | 791 | 524 | 344 | 284 | 139 | 0 | 18 | 56452 |  |  |
| 2000 | 1120 | 85876 | 18307 | 18875 | 42059 | 10892 | 2557 | 584 | 270 | 400 | 316 | 0 | 0 | 191254 |  |  |
| 2001 | 17431 | 65857 | 49449 | 16099 | 25119 | 9486 | 3362 | 1295 | 420 | 134 | 0 | 0 | 0 | 188652 |  |  |
| 2002 | 1113 | 12348 | 10026 | 7112 | 5623 | 5724 | 3616 | 1577 | 448 | 508 | 0 | 0 | 0 | 48095 |  |  |
| 2003 | 60646 | 18043 | 17338 | 21706 | 12578 | 4791 | 3701 | 1424 | 357 | 49 | 0 | 9 | 0 | 140641 |  |  |

Table 6.4.2.4. Catch in number of 1-group (age 1) blue whiting from the spring (March) bottom trawl surveys on the Faroe plateau 1996-2003. The number of 1-group in 2004 was taken out by age readings while for the previous years they were taken out from a visual inspection of the length distributions. There was a clear separation of the 1 and 2group in the data (see Table 6.4.2.5).

| Year | 1-group |
| ---: | ---: |
| 1994 | 1388 |
| 1995 | 1171 |
| 1996 | 4442 |
| 1997 | 1239 |
| 1998 | 262 |
| 1999 | 1108 |
| 2000 | 782 |
| 2001 | 2058 |
| 2002 | 3885 |
| 2003 | 873 |
| 2004 | 13016 |

Table 6.4.2.5. Length distribution (cm) of blue whiting from the spring bottom trawl surveys on the Faroe plateau 1994-2003. Shaded areas in the years 1994-2003 indicate 1-group fish separated from visual inspection of the length distributions (in lack of otolith samples from the catch). In 2004 the 1 -group was separated by age readings from the survey.

| Year |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Length | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $0$ |
| 11 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 12 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 0 | 0 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 14 | 0 | 0 |  |  | 0 | 0 | 0 |  | 0 | 0 | 159 |
| 15 | 0 | 0 |  |  | 0 |  |  |  | 0 | 0 | 1285 |
| 16 |  |  |  |  |  |  |  |  | 0 |  | 3508 |
| 17 |  |  |  |  |  |  |  |  |  |  | 3106 |
| 18 |  |  |  |  |  |  |  |  |  |  | 2035 |
| 19 |  |  |  |  |  |  |  |  |  |  | 1932 |
| 20 |  |  |  | 132 |  |  |  |  |  |  | 1005 |
| 21 |  |  |  | 130 |  |  |  |  |  |  | 589 |
| 22 |  |  |  | 263 | 98 | 222 |  | 1299 |  | 56 | 1544 |
| 23 |  |  | 164 | 331 | 157 | 386 | 105 | 2235 | 1120 | 377 | 1471 |
| 24 | 0 | 112 | 793 | 356 | 183 | 623 | 221 | 2348 | 1047 | 472 | 2401 |
| 25 | 0 | 224 | 1051 | 139 | 179 | 1022 | 312 | 2664 | 1859 | 856 | 2462 |
| 26 | 0 | 291 | 1456 | 94 | 136 | 1259 | 668 | 2687 | 2670 | 602 | 3909 |
| 27 | 8 | 194 | 841 | 82 | 93 | 789 | 1049 | 2493 | 4125 | 691 | 3100 |
| 28 | 4 | 89 | 548 | 61 | 57 | 529 | 1148 | 2954 | 4564 | 240 | 2659 |
| 29 | 5 | 49 | 195 | 93 | 48 | 314 | 1105 | 1774 | 5014 | 184 | 2080 |
| 30 | 23 | 27 | 159 | 16 | 28 | 105 | 939 | 731 | 4852 | 229 | 965 |
| 31 | 36 | 28 | 108 | 38 | 43 | 140 | 549 | 526 | 2924 | 126 | 1311 |
| 32 | 43 | 176 | 65 | 30 | 63 | 114 | 434 | 94 | 1899 | 90 | 754 |
| 33 | 48 | 79 | 112 | 32 | 12 | 96 | 361 | 201 | 1812 | 18 | 452 |
| 34 | 15 | 154 | 159 | 40 | 26 | 61 | 196 | 140 | 928 | 32 | 233 |
| 35 | 24 | 252 | 74 | 14 | 33 | 63 | 172 | 158 | 341 | 9 | 282 |
| 36 | 40 | 134 | 134 | 45 | 20 | 2 | 149 | 101 | 418 | 0 | 22 |
| 37 | 40 | 201 | 127 | 58 | 14 | 50 | 126 | 40 | 253 | 0 | 52 |
| 38 | 31 | 230 | 27 | 41 | 6 | 33 | 47 | 2 | 61 | 0 | 9 |
| 39 | 31 | 107 | 19 | 0 | 8 | 12 | 16 | 0 | 126 | 0 | 0 |
| 40 | 16 | 93 | 29 | 3 | 0 | 0 | 17 | 0 | 0 | 0 | 0 |
| 41 | 7 | 12 | 12 | 0 | 0 | 12 | 15 | 0 | 0 | 0 | 0 |
| 42 | 2 | 0 | 0 | 0 | 0 | 16 | 0 | 0 | 0 | 0 | 0 |
| 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 44 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 45 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $0$ |
| Sum | 1761 | 3623 | 10515 | 3237 | 1466 | 6956 | 8411 | 22505 | 37898 | 4855 | 37325 |
| 1 -group | 1388 | 1171 | 4442 | 1239 | 262 | 1108 | 782 | 2058 | 3885 | 873 | 13016 |

Table 6.4.2.6. Abundance index of blue whiting in the Norwegian shrimp survey in the northern North Sea in October. Blue whiting <19cm in total body length most likely belong to 0-group.


Table 6.4.3.1. Blue whiting. Age stratified Spanish cpue (not used in the assessment)

| Numbers age |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 4 | 5 | 6 | 7 | total |
| 1982 |  |  |  |  |  |  |  |
| 1983 | 7196 | 16392 | 9311 | 7476 | 6326 | 1718 | 48419 |
| 1984 | 13710 | 27286 | 14845 | 4836 | 1755 | 1750 | 64182 |
| 1985 | 14573 | 23823 | 14126 | 6256 | 1232 | 217 | 60227 |
| 1986 | 3721 | 14131 | 14745 | 7113 | 1278 | 505 | 41493 |
| 1987 | 25328 | 13153 | 6664 | 2938 | 1029 | 166 | 49278 |
| 1988 | 7778 | 21473 | 18436 | 6391 | 1300 | 781 | 56159 |
| 1989 | 15272 | 18486 | 17160 | 8374 | 3760 | 1003 | 64055 |
| 1990 | 21444 | 19407 | 5194 | 1803 | 1357 | 451 | 49656 |
| 1991 | 15924 | 15370 | 4989 | 2329 | 1045 | 440 | 40097 |
| 1992 | 10007 | 24235 | 9671 | 4316 | 1194 | 462 | 49885 |
| 1993 | 4036 | 13991 | 22493 | 7979 | 1354 | 658 | 50511 |
| 1994 | 543 | 6066 | 15917 | 7474 | 2990 | 1055 | 34045 |
| 1995 | 9090 | 14409 | 6833 | 4551 | 1990 | 623 | 37496 |
| 1996 | 3905 | 14557 | 14449 | 3931 | 3639 | 1834 | 42315 |
| 1997 | 8742 | 15875 | 11134 | 3698 | 1046 | 450 | 40945 |
| 1998 | 5884 | 13236 | 9803 | 10844 | 5229 | 1153 | 46149 |
| 1999 | 2048 | 10268 | 20242 | 9833 | 6287 | 3047 | 51725 |
| 2000 | 6207 | 15518 | 13987 | 5375 | 1264 | 1414 | 43765 |
| 2001 | 16223 | 16488 | 6830 | 1620 | 1148 | 162 | 42471 |
| 2002 | 10520 | 13725 | 10265 | 3385 | 336 | 69 | 38300 |
| 2003 | 9069 | 10461 | 6517 | 3983 | 1932 | 737 | 32699 |

Table 6.4.4.3.1. Summary results of blue whiting stock assessment by means of ISVPA

|  | R(1) | B(1+) | SSB | F(3-7) |
| :---: | :---: | :---: | :---: | :---: |
| 1981 | 6707.9 | 38985.3 | 4051.4 | 0.242 |
| 1982 | 7874.8 | 32997.4 | 3256.8 | 0.178 |
| 1983 | 4605.2 | 26474.9 | 2466.9 | 0.304 |
| 1984 | 17840.8 | 35131.4 | 1891.8 | 0.364 |
| 1985 | 16083.7 | 40054.9 | 1891.0 | 0.386 |
| 1986 | 19619.3 | 47445.5 | 2381.0 | 0.522 |
| 1987 | 9079.1 | 40100.3 | 2368.5 | 0.512 |
| 1988 | 9951.1 | 35438.8 | 2167.7 | 0.576 |
| 1989 | 10352.5 | 32036.4 | 1920.2 | 0.535 |
| 1990 | 14670.5 | 34759.7 | 1742.0 | 0.615 |
| 1991 | 7374.7 | 29056.1 | 1794.5 | 0.244 |
| 1992 | 6810.8 | 27783.5 | 2146.1 | 0.220 |
| 1993 | 14402.4 | 34379.2 | 2140.7 | 0.198 |
| 1994 | 17494.9 | 42746.0 | 2418.5 | 0.211 |
| 1995 | 7169.7 | 38540.2 | 2795.3 | 0.256 |
| 1996 | 14250.7 | 41552.5 | 2723.1 | 0.316 |
| 1997 | 28163.8 | 56851.2 | 2587.8 | 0.305 |
| 1998 | 32914.6 | 72980.4 | 2959.5 | 0.477 |
| 1999 | 38579.7 | 87789.8 | 3436.8 | 0.433 |
| 2000 | 50376.4 | 110603.6 | 4008.4 | 0.510 |
| 2001 | 70135.7 | 144073.7 | 4996.3 | 0.464 |
| 2002 | 38704.2 | 137333.2 | 6192.4 | 0.650 |
| 2003 | 36608.6 | 122599.3 | 6462.7 | 0.530 |

Table 6.4.4.3.2. Blue whiting. Estimates of stock abundance by means of ISVPA

|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $10+$ |  |  |  |  |  |  |  |  |  |  |
| 1981 | 6707.94 | 4138.11 | 7688.27 | 5674.44 | 2024.87 | 1882.14 | 1677.73 | 2288.92 | 1957.72 | 4945.12 |
| 1982 | 7874.84 | 5165.39 | 3096.57 | 5398.34 | 3741.02 | 1312.84 | 1137.39 | 1016.63 | 1262.54 | 2991.86 |
| 1983 | 4605.22 | 6155.25 | 3951.90 | 2260.22 | 3763.94 | 2577.28 | 860.39 | 746.83 | 625.42 | 928.40 |
| 1984 | 17840.83 | 3497.21 | 4511.28 | 2674.16 | 1411.35 | 2300.26 | 1436.70 | 481.35 | 368.34 | 609.92 |
| 1985 | 16083.68 | 13376.73 | 2514.50 | 2948.77 | 1585.58 | 815.01 | 1184.82 | 743.37 | 212.00 | 590.41 |
| 1986 | 19619.29 | 12004.63 | 9552.19 | 1623.14 | 1715.61 | 896.78 | 407.61 | 595.47 | 313.55 | 717.21 |
| 1987 | 9079.11 | 14267.44 | 8239.50 | 5727.60 | 842.44 | 855.23 | 372.79 | 170.71 | 188.53 | 356.92 |
| 1988 | 9951.13 | 6615.34 | 9821.86 | 4968.47 | 2999.32 | 424.14 | 360.86 | 158.43 | 55.37 | 83.90 |
| 1989 | 10352.47 | 7169.31 | 4475.79 | 5730.26 | 2468.68 | 1424.19 | 163.68 | 140.46 | 44.28 | 67.31 |
| 1990 | 14670.52 | 7512.04 | 4904.27 | 2666.73 | 2944.32 | 1216.93 | 582.08 | 67.41 | 43.25 | 152.17 |
| 1991 | 7374.71 | 10500.79 | 5031.66 | 2806.32 | 1284.17 | 1349.89 | 444.57 | 214.67 | 17.11 | 32.21 |
| 1992 | 6810.83 | 5675.44 | 7850.81 | 3527.45 | 1845.91 | 830.54 | 813.01 | 268.50 | 117.87 | 43.10 |
| 1993 | 14402.38 | 5271.30 | 4279.12 | 5586.11 | 2370.82 | 1222.12 | 516.23 | 506.57 | 153.87 | 70.67 |
| 1994 | 17494.86 | 11203.55 | 4004.43 | 3085.07 | 3826.47 | 1602.30 | 780.81 | 330.53 | 301.25 | 116.69 |
| 1995 | 7169.72 | 13566.97 | 8471.86 | 2863.89 | 2088.89 | 2553.79 | 1006.68 | 491.70 | 192.18 | 134.56 |
| 1996 | 14250.68 | 5502.78 | 10102.40 | 5897.02 | 1864.27 | 1335.79 | 1514.71 | 598.83 | 264.36 | 221.65 |
| 1997 | 28163.76 | 10792.72 | 4016.72 | 6786.37 | 3642.60 | 1125.91 | 732.55 | 833.82 | 288.60 | 468.18 |
| 1998 | 32914.63 | 21379.96 | 7905.92 | 2715.41 | 4231.58 | 2222.64 | 626.29 | 408.96 | 410.00 | 164.97 |
| 1999 | 38579.73 | 24134.90 | 14861.22 | 4854.87 | 1463.09 | 2198.93 | 983.10 | 278.83 | 143.04 | 292.11 |
| 2000 | 50376.44 | 28526.22 | 16990.94 | 9344.88 | 2713.66 | 791.79 | 1032.19 | 464.11 | 106.94 | 256.43 |
| 2001 | 70135.74 | 36715.44 | 19645.56 | 10253.34 | 4899.34 | 1368.02 | 334.75 | 439.53 | 151.03 | 130.91 |
| 2002 | 38704.19 | 51554.62 | 25617.13 | 12149.16 | 5585.42 | 2576.99 | 616.01 | 151.68 | 158.10 | 219.89 |
| 2003 | 36608.63 | 27546.10 | 34228.73 | 14409.23 | 5688.67 | 2481.29 | 895.28 | 216.24 | 35.12 | 490.06 |

Table 6.4.4.3.3. Blue whiting. Estimates of fishing mortality coefficients by means of ISVPA

|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1981 | 0.0613 | 0.0899 | 0.1536 | 0.2166 | 0.2333 | 0.3037 | 0.3010 | 0.3950 | 0.3950 | 0.3950 |
| 1982 | 0.0464 | 0.0678 | 0.1148 | 0.1606 | 0.1726 | 0.2226 | 0.2207 | 0.2858 | 0.2858 | 0.2858 |
| 1983 | 0.0752 | 0.1107 | 0.1906 | 0.2709 | 0.2925 | 0.3844 | 0.3808 | 0.5068 | 0.5068 | 0.5068 |
| 1984 | 0.0880 | 0.1299 | 0.2252 | 0.3227 | 0.3491 | 0.4634 | 0.4589 | 0.6200 | 0.6200 | 0.6200 |
| 1985 | 0.0925 | 0.1368 | 0.2377 | 0.3416 | 0.3699 | 0.4929 | 0.4880 | 0.6632 | 0.6632 | 0.6632 |
| 1986 | 0.1185 | 0.1764 | 0.3115 | 0.4558 | 0.4962 | 0.6778 | 0.6704 | 0.9501 | 0.9501 | 0.9501 |
| 1987 | 0.1166 | 0.1734 | 0.3058 | 0.4469 | 0.4862 | 0.6629 | 0.6557 | 0.9258 | 0.9258 | 0.9258 |
| 1988 | 0.1279 | 0.1907 | 0.3389 | 0.4994 | 0.5448 | 0.7521 | 0.7436 | 1.0747 | 1.0747 | 1.0747 |
| 1989 | 0.1207 | 0.1797 | 0.3178 | 0.4659 | 0.5074 | 0.6948 | 0.6871 | 0.9780 | 0.9780 | 0.9780 |
| 1990 | 0.1344 | 0.2008 | 0.3582 | 0.5307 | 0.5799 | 0.8070 | 0.7975 | 1.1711 | 1.1711 | 1.1711 |
| 1991 | 0.0619 | 0.0908 | 0.1552 | 0.2189 | 0.2358 | 0.3070 | 0.3043 | 0.3995 | 0.3995 | 0.3995 |
| 1992 | 0.0562 | 0.0824 | 0.1403 | 0.1973 | 0.2124 | 0.2755 | 0.2731 | 0.3567 | 0.3567 | 0.3567 |
| 1993 | 0.0512 | 0.0749 | 0.1272 | 0.1783 | 0.1918 | 0.2480 | 0.2459 | 0.3197 | 0.3197 | 0.3197 |
| 1994 | 0.0543 | 0.0795 | 0.1352 | 0.1899 | 0.2044 | 0.2648 | 0.2625 | 0.3422 | 0.3422 | 0.3422 |
| 1995 | 0.0646 | 0.0949 | 0.1623 | 0.2293 | 0.2471 | 0.3224 | 0.3194 | 0.4205 | 0.4205 | 0.4205 |
| 1996 | 0.0779 | 0.1148 | 0.1979 | 0.2818 | 0.3043 | 0.4008 | 0.3970 | 0.5299 | 0.5299 | 0.5299 |
| 1997 | 0.0756 | 0.1113 | 0.1915 | 0.2723 | 0.2940 | 0.3865 | 0.3829 | 0.5099 | 0.5099 | 0.5099 |
| 1998 | 0.1103 | 0.1637 | 0.2876 | 0.4184 | 0.4546 | 0.6157 | 0.6092 | 0.8505 | 0.8505 | 0.8505 |
| 1999 | 0.1019 | 0.1510 | 0.2639 | 0.3817 | 0.4140 | 0.5563 | 0.5506 | 0.7584 | 0.7584 | 0.7584 |
| 2000 | 0.1163 | 0.1730 | 0.3051 | 0.4457 | 0.4849 | 0.6609 | 0.6537 | 0.9226 | 0.9226 | 0.9226 |
| 2001 | 0.1078 | 0.1599 | 0.2806 | 0.4074 | 0.4425 | 0.5979 | 0.5916 | 0.8225 | 0.8225 | 0.8225 |
| 2002 | 0.1401 | 0.2096 | 0.3754 | 0.5588 | 0.6114 | 0.8573 | 0.8469 | 1.2631 | 1.2631 | 1.2631 |
| 2003 | 0.1199 | 0.1785 | 0.3155 | 0.4621 | 0.5031 | 0.6884 | 0.6808 | 0.9675 | 0.9675 | 0.9675 |

Table 6.4.4.4.1. Blue whiting. Description of ICA runs, performed by WG in 2004. Mean F and SSB over the time series presented for each run to aid comparison with values in the final year.

| Run | Separable period | $\begin{gathered} \text { Ages } \\ \text { removed } \end{gathered}$ | Down weighting | No. fleets | Ref. age | Exploitation pattern | Age 1 weight | Mean |  |  | $\begin{gathered} \text { SSB } \\ \text { Mean } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  | F | F | SSB |  |
| 10 | 5 | No | Manual | 3 | 6 | 1 period | 1 | 0.55 | 0.40 | 5.8 | 2.3 |
| 11 | 5 | No | Iterative | 6 | 6 | 1 period | 1 | 0.50 | 0.40 | 6.4 | 2.3 |
| 12 | 5 | No | Manual | 6 | 6 | 1 period | 1 | 0.47 | 0.30 | 6.4 | 2.3 |
| 13 | 8 | No | Manual | 6 | 6 | 1 period | 1 | 0.45 | 0.40 | 5.3 | 2.4 |
| 14 | 4 | No | Manual | 6 | 6 | 1 period | 1 | 0.50 | 0.40 | 6.5 | 2.3 |
| 15 | 5 | No | Manual | 6 | 6 | two periods of 3 | 1 | 0.61 | 0.38 | 5.4 | 2.4 |
| 19 | 8 | No | Manual | 6 | 3 | 1 period | 1 | 0.48 | 0.33 | 4.8 | 2.6 |
| 23 | 5 | No | Manual | 6 | 6 | two periods of 3 | 1 | 0.47 | 0.38 | 6.4 | 2.4 |
| 16b | 5 | No | Manual | 6 | 3 | 1 period | 1 | 0.39 | 0.31 | 6.8 | 2.6 |
| 24 | 5 | No | Manual | 6 | 3 | 1 period | 0.1 | 0.34 | 0.32 | 6.7 | 2.6 |
| 18b | 3 | No | Manual | 6 | 3 | 1 period | 1 | 0.33 | 0.29 | 7.7 | 2.9 |
| 25 | 8 | No | Manual | 6 | 3 | 1 period | 1 | 0.37 | 0.32 | 5.8 | 2.6 |
| 20b | 5 | No | Manual | 6 | 3 | 1 period | 0.5 | 0.39 | 0.32 | 6.7 | 2.6 |
| 21b | 5 | Yes | Manual | 6 | 3 | 1 period | 0.5 | 0.31 | 0.30 | 8.0 | 2.9 |
| 22b | 8 | Yes | Manual | 6 | 3 | 1 period | 0.5 | 0.27 | 0.30 | 7.9 | 3.0 |
| 26* | 8 | No | Manual | 6 | 3 | 1 period | 0.5 | 0.37 | 0.30 | 6.1 | 2.8 |

* Final

Table 6.4.4.4.2. Summary table for final ICA run in 2004 for blue whiting.

| Year | Recruits, age 1 (‘000s) | SSB (t) | Landings (t) | $\mathrm{F} \mathrm{year}^{-1}$ | SOP $\%$ |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |
| 1981 | 3597990 | 4083115 | 907732 | 0.2158 | 98 |
| 1982 | 4544990 | 2931273 | 513203 | 0.1712 | 93 |
| 1983 | 16805190 | 1993806 | 561332 | 0.2009 | 101 |
| 1984 | 19666890 | 1665859 | 626592 | 0.2541 | 101 |
| 1985 | 11068850 | 1919693 | 676812 | 0.317 | 99 |
| 1986 | 9594110 | 2225553 | 801786 | 0.4671 | 94 |
| 1987 | 9282250 | 1899261 | 656588 | 0.3842 | 100 |
| 1988 | 7616070 | 1624370 | 552020 | 0.4682 | 99 |
| 1989 | 10083700 | 1576645 | 598147 | 0.477 | 94 |
| 1990 | 22643940 | 1557809 | 558788 | 0.4033 | 100 |
| 1991 | 9938560 | 2071724 | 363724 | 0.2031 | 99 |
| 1992 | 6625190 | 2717521 | 473789 | 0.1494 | 99 |
| 1993 | 6078760 | 2622049 | 475143 | 0.1696 | 99 |
| 1994 | 6850130 | 2566783 | 458028 | 0.17 | 100 |
| 1995 | 8333210 | 2307944 | 505938 | 0.2134 | 100 |
| 1996 | 22692290 | 2148154 | 629286 | 0.2922 | 101 |
| 1997 | 46643560 | 2321325 | 640089 | 0.2828 | 100 |
| 1998 | 29952190 | 3101413 | 1123732 | 0.4196 | 99 |
| 1999 | 21828850 | 3811625 | 1251463 | 0.3524 | 99 |
| 2000 | 42443380 | 3879908 | 1409143 | 0.4499 | 99 |
| 2001 | 63280810 | 4208597 | 1775305 | 0.454 | 100 |
| 2002 | 47627280 | 5138254 | 1556955 | 0.4497 | 100 |
| 2003 | 62652030 | 6087012 | 2321407 | 0.3691 | 98 |

Table 6.4.4.4.3. Predicted catch numbers for separable period, as estimated in ICA final run for blue whiting in 2004.

| AGE | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |
| 1 | 916.1 | 1823.6 | 1720 | 1058 | 2607.4 | 3921.6 | 2924.2 | 3176.6 |
| 2 | 459.6 | 1205.9 | 3618.1 | 1926.9 | 1788.9 | 3455 | 5101.5 | 3187.2 |
| 3 | 681.5 | 809.9 | 3121.1 | 5310.4 | 4207.1 | 3039.7 | 5766.9 | 7247.5 |
| 4 | 681.2 | 657.9 | 1123.8 | 2406.7 | 6086.2 | 3665.3 | 2600.8 | 4261.7 |
| 5 | 444.2 | 366.5 | 512.4 | 470.3 | 1525.6 | 2868.1 | 1694.5 | 1036.5 |
| 6 | 515.1 | 288.5 | 342.1 | 259.6 | 358.1 | 867.6 | 1600.7 | 818.2 |
| 7 | 600 | 276.6 | 223.8 | 142.4 | 163.7 | 167.5 | 398.1 | 634 |
| 8 | 245 | 423.3 | 277.7 | 122.2 | 116.2 | 99.5 | 100 | 206.8 |
| 9 | 98.4 | 110.5 | 276.2 | 95.9 | 64.3 | 44.9 | 37.7 | 32.7 |

Table 6.4.4.4.4. Fishing mortalities per year, as estimated in ICA final run for blue whiting in 2004.

| AGE | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.0823 | 0.0366 | 0.1619 | 0.1372 | 0.139 | 0.0776 | 0.1047 | 0.0635 | 0.0993 | 0.0816 | 0.03 |
| 2 | 0.0986 | 0.1179 | 0.191 | 0.247 | 0.1746 | 0.121 | 0.0917 | 0.1233 | 0.1451 | 0.1093 | 0.0684 |
| 3 | 0.1562 | 0.1261 | 0.1631 | 0.2308 | 0.3332 | 0.2383 | 0.1509 | 0.133 | 0.352 | 0.1964 | 0.1089 |
| 4 | 0.1086 | 0.1817 | 0.1554 | 0.2333 | 0.2371 | 0.5577 | 0.4063 | 0.2063 | 0.2527 | 0.342 | 0.1268 |
| 5 | 0.2682 | 0.1174 | 0.1426 | 0.2287 | 0.2639 | 0.5478 | 0.4879 | 0.544 | 0.4183 | 0.2624 | 0.2751 |
| 6 | 0.293 | 0.21 | 0.1923 | 0.3253 | 0.4343 | 0.4554 | 0.3732 | 1.0442 | 0.6518 | 0.573 | 0.201 |
| 7 | 0.253 | 0.2208 | 0.351 | 0.2522 | 0.3164 | 0.5365 | 0.5029 | 0.4133 | 0.7103 | 0.6426 | 0.3035 |
| 8 | 0.3071 | 0.2298 | 0.3909 | 0.4231 | 0.2648 | 0.4301 | 0.7525 | 0.5156 | 0.4649 | 1.2155 | 0.2025 |
| 9 | 0.2521 | 0.2193 | 0.302 | 0.3778 | 0.3714 | 0.4607 | 0.4178 | 0.4613 | 0.4924 | 0.4946 | 0.212 |
| 10 | 0.2521 | 0.2193 | 0.302 | 0.3778 | 0.3714 | 0.4607 | 0.4178 | 0.4613 | 0.4924 | 0.4946 | 0.212 |

Table 6.4.4.4.4. (continued). Fishing mortalities per year, as estimated in ICA final run for blue whiting in 2004.

| AGE | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | 0.0702 | 0.0489 | 0.0507 | 0.04 | 0.0455 | 0.0441 | 0.0654 | 0.0549 | 0.0701 | 0.0707 | 0.0701 | 0.0575 |
| 2 | 0.0956 | 0.0688 | 0.0254 | 0.076 | 0.0804 | 0.0778 | 0.1154 | 0.097 | 0.1238 | 0.1249 | 0.1237 | 0.1015 |
| 3 | 0.1489 | 0.1237 | 0.1108 | 0.1309 | 0.2049 | 0.1983 | 0.2942 | 0.2471 | 0.3154 | 0.3183 | 0.3152 | 0.2588 |
| 4 | 0.1689 | 0.2076 | 0.1063 | 0.1993 | 0.3218 | 0.3114 | 0.4621 | 0.3881 | 0.4954 | 0.4999 | 0.4951 | 0.4064 |
| 5 | 0.1267 | 0.1773 | 0.2471 | 0.2438 | 0.2967 | 0.2871 | 0.426 | 0.3578 | 0.4568 | 0.4609 | 0.4565 | 0.3747 |
| 6 | 0.1795 | 0.1568 | 0.1766 | 0.2539 | 0.3307 | 0.32 | 0.4748 | 0.3988 | 0.5091 | 0.5138 | 0.5088 | 0.4177 |
| 7 | 0.1228 | 0.1825 | 0.2092 | 0.2393 | 0.3071 | 0.2972 | 0.4411 | 0.3704 | 0.4729 | 0.4772 | 0.4726 | 0.388 |
| 8 | 0.2446 | 0.0992 | 0.2312 | 0.3014 | 0.3826 | 0.3703 | 0.5495 | 0.4615 | 0.5892 | 0.5945 | 0.5888 | 0.4833 |
| 9 | 0.1954 | 0.1769 | 0.1686 | 0.2353 | 0.3073 | 0.2974 | 0.4413 | 0.3706 | 0.4731 | 0.4774 | 0.4729 | 0.3881 |
| 10 | 0.1954 | 0.1769 | 0.1686 | 0.2353 | 0.3073 | 0.2974 | 0.4413 | 0.3706 | 0.4731 | 0.4774 | 0.4729 | 0.3881 |

Table 6.4.4.5.1 Blue whiting. Objective function contributions (negative log-likelihood) for the

| Data source | Likelihood <br> contribution |
| :--- | ---: |
| Catch data | -149.05 |
| Tuning data |  |
| Norwegian acoustic survey, 1981-1990 | 2.87 |
| 1991-2003 | 5.05 |
| Russian acoustic survey1982-1991 | 2.08 |
| 1992-1996 | 1.38 |
| Norwegian Sea 1981-1990 | 15.42 |
| 1991-2001 | 16.55 |
| Total tuning data | 59.26 |

Table 6.4.4.5.2 Blue whiting, Survey catachability by fleet and period as estimated by SMS

| Survey | Period | age 2 | age 3 | age 4 | age 5 | age 6 | age 7-8 |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Norwegian acoustic survey | $1981-1990$ | 0.78 | 1.98 | 2.61 | 2.96 | 2.80 | 3.04 |
|  | $1991-2003$ | 0.95 | 2.19 | 3.00 | 1.98 | 1.59 | 1.53 |


| Survey | Period | age 3 | age 4 | age 5 | age 6 | age 7-8 |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: |
| Russian acoustic survey | $1982-1991$ | 1.56 | 2.18 | 2.59 | 3.19 | 3.85 |
|  | $1992-1996$ | 2.69 | 3.35 | 2.80 | 1.60 | 1.26 |


| Survey | Period | age 1 | age 2 | age 3 | age 4 | age 5 | age 6-7 |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Norwegian Sea | $1981-1990$ | 0.40 | 0.45 | 0.68 | 0.82 | 0.82 | 0.69 |
|  | $1991-2001$ | 0.64 | 0.40 | 0.56 | 0.47 | 0.27 | 0.22 |

Table 6.4.4.5.3 Blue Whiting, SMS exploratory run, Stock numbers (thousands)

| Age |  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 | 3107287 | 3993368 | 15287046 | 19734898 | 11006136 | 8756361 | 9632978 | 6952793 |
|  | 2 | 3419737 | 2357016 | 3072474 | 11614377 | 14675308 | 8118077 | 6319045 | 7165759 |
|  | 3 | 4416428 | 2489815 | 1753948 | 2242461 | 8201771 | 10234512 | 5474521 | 4464704 |
|  | 4 | 2440933 | 3038455 | 1769304 | 1211081 | 1474498 | 5293882 | 6285043 | 3602515 |
|  | 5 | 2206883 | 1599287 | 2075021 | 1164635 | 748793 | 890266 | 2998842 | 3889850 |
|  | 6 | 2135198 | 1379440 | 1051120 | 1304335 | 678596 | 423927 | 466552 | 1749454 |
|  | 7 | 1915789 | 1216853 | 840946 | 603576 | 676466 | 338609 | 190698 | 242362 |
|  | 8 | 1738816 | 1021826 | 702884 | 452560 | 287959 | 308326 | 136520 | 91156 |
|  | 9 | 1501514 | 918893 | 585803 | 374850 | 213408 | 129599 | 122424 | 64504 |
|  | 10 | 3138668 | 2452146 | 1932585 | 1343064 | 810097 | 460640 | 234361 | 168577 |
|  |  |  |  |  |  |  |  |  |  |
| TSB |  | 3208931 | 2569775 | 2049201 | 2191103 | 2720337 | 2913365 | 2583034 | 2385492 |
| SSB |  | 2903101 | 2320272 | 1801178 | 1560608 | 1884231 | 2244204 | 1998779 | 1850030 |
|  |  |  |  |  |  |  |  |  |  |
|  |  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
|  |  | 9 | 970161 | 23335785 | 8257778 | 5574977 | 5244130 | 5873522 | 8021319 |
|  | 2 | 5181629 | 6679984 | 16895114 | 6377019 | 4321040 | 4137598 | 4655447 | 6292154 |
|  | 3 | 5077399 | 3486392 | 4527276 | 12643956 | 4799372 | 3343691 | 3224170 | 3570703 |
|  | 4 | 2950445 | 3107599 | 2156911 | 3244375 | 9136958 | 3427023 | 2428339 | 2253342 |
|  | 5 | 2241717 | 1664179 | 1777140 | 1488975 | 2263750 | 6059402 | 2332750 | 1558030 |
|  | 6 | 2284172 | 1168692 | 882202 | 1183377 | 1004412 | 1487553 | 4091568 | 1480898 |
|  | 7 | 916830 | 1020351 | 533889 | 547333 | 747058 | 643870 | 982891 | 2524064 |
|  | 8 | 117064 | 366603 | 418954 | 314853 | 329485 | 481576 | 427519 | 610268 |
|  | 9 | 43530 | 46090 | 148300 | 245328 | 188282 | 203766 | 308344 | 253004 |
|  | 10 | 111305 | 60962 | 43305 | 112199 | 213801 | 248663 | 289682 | 353910 |

Table 6.4.4.5.4 Blue Whiting, SMS exploratory run, Fishing mortality

| Age |  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 0.0764 | 0.0622 | 0.0748 | 0.0962 | 0.1044 | 0.1262 | 0.0959 | 0.0940 |
|  | 2 | 0.1174 | 0.0955 | 0.1149 | 0.1479 | 0.1604 | 0.1940 | 0.1474 | 0.1445 |
|  | 3 | 0.1740 | 0.1416 | 0.1704 | 0.2193 | 0.2378 | 0.2876 | 0.2185 | 0.2142 |
|  | 4 | 0.2228 | 0.1814 | 0.2182 | 0.2808 | 0.3046 | 0.3683 | 0.2798 | 0.2744 |
|  | 5 | 0.2699 | 0.2197 | 0.2643 | 0.3401 | 0.3689 | 0.4462 | 0.3389 | 0.3324 |
|  | 6 | 0.3623 | 0.2949 | 0.3547 | 0.4566 | 0.4952 | 0.5989 | 0.4549 | 0.4461 |
|  | 7 | 0.4285 | 0.3488 | 0.4196 | 0.5401 | 0.5857 | 0.7084 | 0.5381 | 0.5277 |
|  | 8 | 0.4378 | 0.3564 | 0.4287 | 0.5517 | 0.5984 | 0.7237 | 0.5497 | 0.5391 |
|  | 9 | 0.4378 | 0.3564 | 0.4287 | 0.5517 | 0.5984 | 0.7237 | 0.5497 | 0.5391 |
|  | 10 | 0.4378 | 0.3564 | 0.4287 | 0.5517 | 0.5984 | 0.7237 | 0.5497 | 0.5391 |
| Avg. F | 3-7 | 0.292 | 0.237 | 0.285 | 0.367 | 0.398 | 0.482 | 0.366 | 0.359 |
| Age |  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
|  | 1 | 0.1277 | 0.1230 | 0.0585 | 0.0548 | 0.0370 | 0.0324 | 0.0428 | 0.0583 |
|  | 2 | 0.1963 | 0.1890 | 0.0898 | 0.0842 | 0.0564 | 0.0494 | 0.0653 | 0.0888 |
|  | 3 | 0.2909 | 0.2802 | 0.1332 | 0.1249 | 0.1368 | 0.1199 | 0.1583 | 0.2154 |
|  | 4 | 0.3726 | 0.3588 | 0.1706 | 0.1599 | 0.2107 | 0.1846 | 0.2438 | 0.3318 |
|  | 5 | 0.4514 | 0.4347 | 0.2066 | 0.1937 | 0.2199 | 0.1927 | 0.2544 | 0.3463 |
|  | 6 | 0.6059 | 0.5835 | 0.2774 | 0.2600 | 0.2447 | 0.2144 | 0.2831 | 0.3853 |
|  | 7 | 0.7166 | 0.6901 | 0.3281 | 0.3075 | 0.2391 | 0.2095 | 0.2766 | 0.3765 |
|  | 8 | 0.7321 | 0.7050 | 0.3352 | 0.3142 | 0.2806 | 0.2458 | 0.3246 | 0.4418 |
|  | 9 | 0.7321 | 0.7050 | 0.3352 | 0.3142 | 0.2806 | 0.2458 | 0.3246 | 0.4418 |
|  | 10 | 0.7321 | 0.7050 | 0.3352 | 0.3142 | 0.2806 | 0.2458 | 0.3246 | 0.4418 |
| Avg. F | 3-7 | 0.487 | 0.469 | 0.223 | 0.209 | 0.210 | 0.184 | 0.243 | 0.331 |
| Age |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |  |
|  | 1 | 0.0586 | 0.0819 | 0.0706 | 0.0891 | 0.0836 | 0.0813 | 0.1047 |  |
|  | 2 | 0.0894 | 0.1250 | 0.1077 | 0.1359 | 0.1274 | 0.1240 | 0.1596 |  |
|  | 3 | 0.2167 | 0.3030 | 0.2612 | 0.3294 | 0.3090 | 0.3007 | 0.3871 |  |
|  | 4 | 0.3338 | 0.4667 | 0.4024 | 0.5074 | 0.4759 | 0.4632 | 0.5962 |  |
|  | 5 | 0.3483 | 0.4870 | 0.4199 | 0.5295 | 0.4967 | 0.4834 | 0.6222 |  |
|  | 6 | 0.3876 | 0.5419 | 0.4672 | 0.5892 | 0.5526 | 0.5378 | 0.6923 |  |
|  | 7 | 0.3787 | 0.5295 | 0.4566 | 0.5757 | 0.5400 | 0.5256 | 0.6765 |  |
|  | 8 | 0.4444 | 0.6214 | 0.5358 | 0.6756 | 0.6337 | 0.6168 | 0.7939 |  |
|  | 9 | 0.4444 | 0.6214 | 0.5358 | 0.6756 | 0.6337 | 0.6168 | 0.7939 |  |
|  | 10 | 0.4444 | 0.6214 | 0.5358 | 0.6756 | 0.6337 | 0.6168 | 0.7939 |  |
| Avg. F | 3-7 | 0.333 | 0.466 | 0.401 | 0.506 | 0.475 | 0.462 | 0.595 |  |

Table 6.4.4.5.5 Blue Whiting, SMS exploratory run, Stock summary

| Year | Recruits <br> age 1 | SSB <br> tonnes | $F$ |
| :--- | ---: | ---: | ---: |
| 1981 | 3107287 | 2903101 | 0.292 |
| 1982 | 3993368 | 2320272 | 0.237 |
| 1983 | 15287046 | 1801178 | 0.285 |
| 1984 | 19734898 | 1560608 | 0.367 |
| 1985 | 11006136 | 1884231 | 0.398 |
| 1986 | 8756361 | 2244204 | 0.482 |
| 1987 | 9632978 | 1998779 | 0.366 |
| 1988 | 6952793 | 1850030 | 0.359 |
| 1989 | 9270161 | 1771899 | 0.487 |
| 1990 | 23335785 | 1508598 | 0.469 |
| 1991 | 8257778 | 1955697 | 0.223 |
| 1992 | 5574977 | 2577452 | 0.209 |
| 1993 | 5244130 | 2473462 | 0.21 |
| 1994 | 5873522 | 2363735 | 0.184 |
| 1995 | 8021319 | 2183796 | 0.243 |
| 1996 | 2449664 | 2009328 | 0.331 |
| 1997 | 45328501 | 2038728 | 0.333 |
| 1998 | 27570442 | 3176026 | 0.466 |
| 1999 | 20732646 | 3868681 | 0.401 |
| 2000 | 35451807 | 3876646 | 0.506 |
| 2001 | 56872588 | 3972157 | 0.475 |
| 2002 | 33706885 | 4793498 | 0.462 |
| 2003 | 41437715 | 5523205 | 0.595 |

Table 6.4.4.6.2.1. Summary of recruitment indices for blue whiting.

|  | Survey | Year | Age | Raw numbers | Age-specific |  | Average per age |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | Normalized index | Normalized log(index) | Normalized index | Normalized log(index) |
|  | Icelandic summer acoustic | 2000 | 0 | 10683 | -0.03 | 0.30 |  |  |
|  | North Sea shrimp | 2000 | 0 | 4.233 | -0.39 | 0.39 |  |  |
|  | Faroese bottom trawl summer | 2000 | 0 | 11120 | -0.18 | 0.30 | -0.20 | 0.33 |
|  | Norwegian Sea summer acoustic | 2001 | 1 | 61470 | 2.90 | 1.33 |  |  |
|  | Spanish CPUE | 2001 | 1 | 16223 | 0.97 | 0.85 |  |  |
|  | Icelandic summer acoustic | 2001 | 1 | 4090 | -0.78 | -0.54 |  |  |
|  | Barents Sea bottom trawl | 2001 | 1 | 340768 | 3.92 | 1.52 |  |  |
|  | PGSPFN May acoustic | 2001 | 1 | 85772 | 1.33 | 1.10 |  |  |
|  | Faroese bottom trawl spring | 2001 | 1 | 2058 | -0.19 | 0.24 |  |  |
|  | Faroese bottom trawl summer | 2001 | 1 | 65857 | 1.21 | 1.20 | 1.34 | 0.81 |
|  | Norwegian spawning st. acoustic | 2002 | 2 | 71996 | 3.47 | 1.72 |  |  |
|  | Spanish CPUE | 2002 | 2 | 13725 | -0.48 | -0.34 |  |  |
|  | Icelandic summer acoustic | 2002 | 2 | 3107 | 0.00 | 0.27 |  |  |
|  | PGSPFN May acoustic | 2002 | 2 | 46656 | 1.26 | 0.88 |  |  |
|  | Faroese bottom trawl summer | 2002 | 2 | 10026 | -0.56 | -0.38 | 0.74 | 0.43 |
|  | Icelandic summer acoustic | 2001 | 0 | 27305 | 1.79 | 1.49 |  |  |
|  | North Sea shrimp | 2001 | 0 | 2.735 | -0.43 | 0.24 |  |  |
|  | Faroese bottom trawl summer | 2001 | 0 | 17431 | 0.15 | 0.66 | 0.50 | 0.80 |
|  | Spanish CPUE | 2002 | 1 | 10520 | 0.07 | 0.35 |  |  |
|  | Icelandic summer acoustic | 2002 | 1 | 10785 | 1.05 | 0.87 |  |  |
|  | Barents Sea bottom trawl | 2002 | 1 | 70859 | 0.49 | 1.07 |  |  |
|  | PGSPFN May acoustic | 2002 | 1 | 15251 | -1.05 | -1.30 |  |  |
|  | Faroese bottom trawl spring | 2002 | 1 | 3885 | 0.31 | 0.85 |  |  |
|  | Faroese bottom trawl summer | 2002 | 1 | 12348 | -0.64 | -0.58 | 0.04 | 0.21 |
|  | Norwegian spawning st. acoustic | 2003 | 2 | 23992 | 0.59 | 0.89 |  |  |
|  | Spanish CPUE | 2003 | 2 | 10461 | -1.13 | -1.16 |  |  |
|  | Icelandic summer acoustic | 2003 | 2 | 4899 | 0.91 | 0.86 |  |  |
|  | PGSPFN May acoustic | 2003 | 2 | 21487 | -0.15 | 0.21 |  |  |
|  | Faroese bottom trawl summer | 2003 | 2 | 17338 | -0.06 | 0.28 | 0.03 | 0.22 |
|  | Icelandic summer acoustic | 2002 | 0 | 3815 | -0.79 | -1.00 |  |  |
|  | North Sea shrimp | 2002 | 0 | 0.036 | -0.49 | -1.25 |  |  |
|  | Faroese bottom trawl summer | 2002 | 0 | 1113 | -0.69 | -1.55 | -0.66 | -1.27 |
|  | Spanish CPUE | 2003 | 1 | 9069 | -0.15 | 0.18 |  |  |
|  | Icelandic summer acoustic | 2003 | 1 | 9158 | 0.60 | 0.63 |  |  |
|  | Barents Sea bottom trawl | 2003 | 1 | 7377 | -0.32 | 0.42 |  |  |
|  | PGSPFN May acoustic | 2003 | 1 | 35688 | -0.36 | -0.12 |  |  |
|  | Faroese bottom trawl spring | 2003 | 1 | 873 | -0.51 | -0.59 |  |  |
|  | Faroese bottom trawl summer | 2003 | 1 | 18043 | -0.44 | -0.18 | -0.20 | 0.06 |
|  | Norwegian spawning st. acoustic | 2004 | 2 | 18569 | 0.27 | 0.70 | 0.27 | 0.70 |
|  | Icelandic summer acoustic | 2003 | 0 | 5011 | -0.66 | -0.66 |  |  |
|  | North Sea shrimp | 2003 | 0 | 6.23 | -0.35 | 0.53 |  |  |
|  | Faroese bottom trawl summer | 2003 | 0 | 60646 | 2.37 | 1.67 | 0.46 | 0.51 |
|  | Faroese bottom trawl spring | 2004 | 1 | 13016 | 2.82 | 2.02 | 2.82 | 2.02 |

Table 6.4.5.1.1 Tuning data for the blue whiting assessment. Inside the framed areas constant selection pattern is assumed. $-1=$ missing data.

## 

Norwegian acoustic spawning stock survey, ages 2-8

|  |  |  |  | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 1 | 1 | 1 | 2372 | 7583 | 3253 | 3647 | 4611 | 4638 | 3654 |
| 1982 | 1 | 1 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1983 | 1 | 1 | 1 | 297 | 2108 | 2723 | 6511 | 3735 | 3650 | 3153 |
| 1984 | 1 | 1 | 1 | 15767 | 1721 | 1616 | 1719 | 1858 | 1128 | 567 |
| 1985 | 1 | 1 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1986 | 1 | 1 | 1 | 1003 | 5829 | 4122 | 624 | 228 | 203 | 250 |
| 1987 | 1 | 1 | 1 | 4960 | 8417 | 22589 | 4735 | 282 | 417 | 385 |
| 1988 | 1 | 1 | 1 | 9712 | 9090 | 12367 | 20392 | 7355 | 723 | 599 |
| 1989 | 1 | 1 | 1 | 6787 | 22270 | 9973 | 10504 | 7803 | 933 | 293 |
| 1990 | 1 | 1 | 1 | 14169 | 12670 | 11228 | 5587 | 6556 | 3273 | 516 |
| 1991 | 1 | 1 | 1 | 11147 | 6340 | 8497 | 7407 | 4558 | 2019 | 545 |
| 1992 | 1 | 1 | 1 | 1232 | 26123 | 4719 | 1574 | 1386 | 810 | 616 |
| 1993 | 1 | 1 | 1 | 4489 | 3321 | 26771 | 2643 | 1270 | 557 | 426 |
| 1994 | 1 | 1 | 1 | 1603 | 2950 | 4476 | 11354 | 1742 | 1687 | 908 |
| 1995 | 1 | 1 | 1 | 8538 | 9874 | 7906 | 6861 | 9467 | 1795 | 1083 |
| 1996 | 1 | 1 | 1 | 8781 | 7433 | 8371 | 2399 | 4455 | 4111 | 1202 |
| 1997 | 1 | 1 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1998 | 1 | 1 | 1 | 18218 | 34991 | 4697 | 1674 | 279 | 407 | 381 |
| 1999 | 1 | 1 | 1 | 19034 | 60309 | 26103 | 1481 | 316 | 72 | 153 |
| 2000 | 1 | 1 | 1 | 8613 | 31011 | 41382 | 6843 | 898 | 427 | 228 |
| 2001 | 1 | 1 | 1 | 44162 | 12843 | 13805 | 8292 | 718 | 175 | 51 |
| 2002 | 1 | 1 | 1 | 71996 | 54740 | 12757 | 5266 | 8404 | 1450 | 305 |
| 2003 | 1 | 1 | 1 | 23992 | 70303 | 28756 | 5735 | 2430 | 1708 | 260 |
| 2004 | 1 | 1 | 1 | 18569 | 40669 | 50137 | 15649 | 4454 | 2218 | 1313 |

Russian acoustic spawning stock survey, ages 3-8

|  |  |  |  | 3 | 4 | 5 | 6 | 7 | 8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 1 | 2 | 10 | 540 | 2750 | 1340 | 1380 | 1570 | 2350 |
| 1983 | 1 | 2 | 10 | 2330 | 2930 | 9390 | 3880 | 1970 | 1370 |
| 1984 | 1 | 2 | 10 | 2900 | 800 | 1100 | 4200 | 2200 | 1200 |
| 1985 | 1 | 2 | 10 | 13220 | 930 | 580 | 1780 | 860 | 610 |
| 1986 | 1 | 2 | 10 | 18750 | 23180 | 2540 | 610 | 620 | 750 |
| 1987 | 1 | 2 | 10 | 4480 | 19170 | 5860 | 1070 | 500 | 810 |
| 1988 | 1 | 2 | 10 | 3710 | 4550 | 8610 | 4130 | 1270 | 480 |
| 1989 | 1 | 2 | 10 | 11910 | 7120 | 6670 | 6970 | 4580 | 2750 |
| 1990 | 1 | 2 | 10 | 9740 | 12140 | 5740 | 2580 | 1470 | 220 |
| 1991 | 1 | 2 | 10 | 10300 | 5350 | 5130 | 2630 | 1770 | 870 |
| 1992 | 1 | 2 | 1 | 20010 | 6700 | 1350 | 440 | 390 | 170 |
| 1993 | 1 | 2 | 1 | 4728 | 12337 | 5304 | 2249 | 1316 | 621 |
| 1994 | 1 | 2 | 1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1995 | 1 | 2 | 1 | 12657 | 10028 | 8942 | 2651 | 1093 | 408 |
| 1996 | 1 | 2 | 1 | 15285 | 10629 | 4897 | 6940 | 1482 | 653 |

Norwegian Sea acoustic survey, ages 1-7

|  |  |  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 3 | 3 | 1 | 182 | 728 | 4542 | 3874 | 2678 | 2834 | 2964 |
| 1982 | 3 | 3 | 1 | 184 | 460 | 1242 | 4715 | 3611 | 3128 | 2323 |
| 1983 | 3 | 3 | 1 | 22356 | 396 | 468 | 756 | 1404 | 576 | 468 |
| 1984 | 3 | 3 | 1 | 30380 | 13916 | 833 | 392 | 539 | 539 | 343 |
| 1985 | 3 | 3 | 1 | 5969 | 23876 | 12502 | 658 | 423 | 188 | 235 |
| 1986 | 3 | 3 | 1 | 2324 | 2380 | 7224 | 6944 | 1876 | 952 | 336 |
| 1987 | 3 | 3 | 1 | 8204 | 4032 | 5180 | 5572 | 1204 | 224 | 168 |
| 1988 | 3 | 3 | 1 | 4992 | 2880 | 2640 | 3480 | 912 | 120 | 96 |
| 1989 | 3 | 3 | 1 | 1172 | 1125 | 812 | 379 | 410 | 212 | 22 |
| 1990 | 3 | 3 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1991 | 3 | 3 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1992 | 3 | 3 | 1 | 792 | 1134 | 6939 | 766 | 247 | 172 | 90 |
| 1993 | 3 | 3 | 1 | 830 | 125 | 1070 | 6392 | 1222 | 489 | 248 |
| 1994 | 3 | 3 | 1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1995 | 3 | 3 | 1 | 6974 | 2811 | 1999 | 1209 | 1622 | 775 | 173 |
| 1996 | 3 | 3 | 1 | 23464 | 1057 | 899 | 649 | 436 | 505 | 755 |
| 1997 | 3 | 3 | 1 | 30227 | 25638 | 1524 | 779 | 300 | 407 | 260 |
| 1998 | 3 | 3 | 1 | 24244 | 47815 | 16282 | 556 | 212 | 100 | 64 |
| 1999 | 3 | 3 | 1 | 14367 | 9750 | 23701 | 9754 | 1733 | 466 | 79 |
| 2000 | 3 | 3 | 1 | 25813 | 3298 | 2721 | 3078 | 23 | 46 | 6 |
| 2001 | 3 | 3 | 1 | 61470 | 22051 | 7883 | 3225 | 1824 | 156 | 12 |

Table 6.4.5.2.2 Blue Whiting: AMCI: Fishing mortality at age

| Age | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |
| 1 | 0.079 | 0.051 | 0.150 | 0.167 | 0.153 | 0.141 | 0.117 | 0.082 |
| 2 | 0.105 | 0.084 | 0.111 | 0.164 | 0.178 | 0.226 | 0.196 | 0.171 |
| 3 | 0.192 | 0.145 | 0.172 | 0.223 | 0.261 | 0.340 | 0.297 | 0.256 |
| 4 | 0.229 | 0.177 | 0.207 | 0.264 | 0.284 | 0.407 | 0.384 | 0.336 |
| 5 | 0.270 | 0.201 | 0.230 | 0.295 | 0.317 | 0.441 | 0.415 | 0.394 |
| 6 | 0.426 | 0.315 | 0.366 | 0.467 | 0.515 | 0.675 | 0.598 | 0.571 |
| 7 | 0.426 | 0.315 | 0.366 | 0.467 | 0.515 | 0.675 | 0.598 | 0.571 |
| 8 | 0.426 | 0.315 | 0.366 | 0.467 | 0.515 | 0.675 | 0.598 | 0.571 |
| 9 | 0.426 | 0.315 | 0.366 | 0.467 | 0.515 | 0.675 | 0.598 | 0.571 |
| $10+$ | 0.426 | 0.315 | 0.366 | 0.467 | 0.515 | 0.675 | 0.598 | 0.571 |
| $F(3-7)$ | 0.309 | 0.230 | 0.268 | 0.343 | 0.378 | 0.508 | 0.459 | 0.425 |
| Age | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| 0 |  |  |  |  |  |  |  |  |
| 1 | 0.103 | 0.093 | 0.035 | 0.056 | 0.052 | 0.051 | 0.053 | 0.087 |
| 2 | 0.197 | 0.188 | 0.077 | 0.082 | 0.076 | 0.068 | 0.082 | 0.108 |
| 3 | 0.310 | 0.298 | 0.124 | 0.130 | 0.123 | 0.117 | 0.143 | 0.190 |
| 4 | 0.395 | 0.387 | 0.159 | 0.168 | 0.162 | 0.150 | 0.186 | 0.250 |
| 5 | 0.480 | 0.473 | 0.204 | 0.204 | 0.193 | 0.189 | 0.235 | 0.302 |
| 6 | 0.676 | 0.694 | 0.298 | 0.294 | 0.268 | 0.252 | 0.300 | 0.393 |
| 7 | 0.676 | 0.694 | 0.298 | 0.294 | 0.268 | 0.252 | 0.300 | 0.393 |
| 8 | 0.676 | 0.694 | 0.298 | 0.294 | 0.268 | 0.252 | 0.300 | 0.393 |
| 9 | 0.676 | 0.694 | 0.298 | 0.294 | 0.268 | 0.252 | 0.300 | 0.393 |
| 10+ | 0.676 | 0.694 | 0.298 | 0.294 | 0.268 | 0.252 | 0.300 | 0.393 |
| F(3-7) | 0.507 | 0.509 | 0.217 | 0.218 | 0.202 | 0.192 | 0.233 | 0.306 |
| Age | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |  |
| 0 |  |  |  |  |  |  |  |  |
| 1 | 0.074 | 0.084 | 0.061 | 0.066 | 0.090 | 0.076 | 0.147 |  |
| 2 | 0.112 | 0.155 | 0.137 | 0.158 | 0.179 | 0.176 | 0.188 |  |
| 3 | 0.202 | 0.292 | 0.274 | 0.326 | 0.365 | 0.360 | 0.392 |  |
| 4 | 0.260 | 0.372 | 0.376 | 0.457 | 0.519 | 0.533 | 0.581 |  |
| 5 | 0.306 | 0.412 | 0.373 | 0.459 | 0.513 | 0.531 | 0.624 |  |
| 6 | 0.396 | 0.545 | 0.481 | 0.606 | 0.651 | 0.672 | 0.727 |  |
| 7 | 0.396 | 0.545 | 0.481 | 0.606 | 0.651 | 0.672 | 0.727 |  |
| 8 | 0.396 | 0.545 | 0.481 | 0.606 | 0.651 | 0.672 | 0.727 |  |
| 9 | 0.396 | 0.545 | 0.481 | 0.606 | 0.651 | 0.672 | 0.727 |  |
| 10+ | 0.396 | 0.545 | 0.481 | 0.606 | 0.651 | 0.672 | 0.727 |  |
| $F(3-7)$ | 0.312 | 0.433 | 0.397 | 0.491 | 0.540 | 0.554 | 0.610 |  |

Table 6.4.5.2.3 Blue Whiting: AMCI: Stock numbers at age

| Age |  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 |  |  |  |  |  |  |  |  |
|  | 1 | 3,358,560 | 3,920,662 | 11,450,733 | 19,016,674 | 12,011,291 | 9,429,937 | 9,260,073 | 7,706,866 |
|  | 2 | 3,646,770 | 2,540,889 | 3,049,799 | 8,071,385 | 13,176,051 | 8,441,022 | 6,703,938 | 6,744,415 |
|  | 3 | 4,653,650 | 2,688,114 | 1,912,560 | 2,233,705 | 5,609,587 | 9,026,187 | 5,514,923 | 4,510,929 |
|  | 4 | 2,972,687 | 3,144,862 | 1,903,280 | 1,319,077 | 1,462,819 | 3,536,511 | 5,259,410 | 3,355,998 |
|  | 5 | 2,382,764 | 1,935,421 | 2,156,270 | 1,266,828 | 829,411 | 901,975 | 1,926,442 | 2,931,706 |
|  | 6 | 2,318,133 | 1,489,750 | 1,296,697 | 1,402,144 | 772,294 | 494,748 | 475,267 | 1,041,651 |
|  | 7 | 1,895,226 | 1,239,011 | 890,443 | 736,471 | 719,817 | 377,893 | 206,160 | 213,916 |
|  | 8 | 1,786,152 | 1,012,973 | 740,573 | 505,735 | 378,081 | 352,216 | 157,467 | 92,792 |
|  | 9 | 1,490,579 | 954,674 | 605,467 | 420,615 | 259,629 | 185,000 | 146,767 | 70,875 |
|  | 10+ | 3,125,753 | 2,467,369 | 2,045,399 | 1,505,584 | 988,851 | 610,897 | 331,648 | 215,333 |
| Age |  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
|  | 0 |  |  |  |  |  |  |  |  |
|  | 1 | 9,015,810 | 21,516,302 | 9,083,595 | 6,104,599 | 5,532,517 | 6,497,652 | 7,573,674 | 20,506,834 |
|  | 2 | 5,810,531 | 6,657,906 | 16,054,254 | 7,181,951 | 4,726,460 | 4,298,298 | 5,054,060 | 5,881,274 |
|  | 3 | 4,654,779 | 3,906,181 | 4,515,464 | 12,168,161 | 5,417,764 | 3,587,060 | 3,288,330 | 3,811,197 |
|  | 4 | 2,859,083 | 2,794,777 | 2,374,004 | 3,266,477 | 8,743,640 | 3,924,255 | 2,612,946 | 2,333,582 |
|  | 5 | 1,963,991 | 1,577,184 | 1,553,519 | 1,658,442 | 2,261,142 | 6,089,206 | 2,764,650 | 1,776,401 |
|  | 6 | 1,618,655 | 994,703 | 804,571 | 1,037,457 | 1,106,862 | 1,526,552 | 4,124,997 | 1,789,768 |
|  | 7 | 481,955 | 674,164 | 406,848 | 488,895 | 632,802 | 693,470 | 971,525 | 2,501,784 |
|  | 8 | 98,975 | 200,732 | 275,743 | 247,220 | 298,204 | 396,463 | 441,337 | 589,224 |
|  | 9 | 42,933 | 41,223 | 82,102 | 167,554 | 150,793 | 186,830 | 252,316 | 267,668 |
|  | 10+ | 132,424 | 73,035 | 46,733 | 78,287 | 149,952 | 188,422 | 238,818 | 297,869 |
| Age |  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 |
|  | 0 |  |  |  |  |  |  |  |  |
|  | 1 | 41,591,472 | 28,588,844 | 19,033,352 | 33,972,100 | 47,003,316 | 34,434,780 | 20,000,000 | 20,000,000 |
|  | 2 | 15,383,627 | 31,619,636 | 21,525,614 | 14,655,518 | 26,035,112 | 35,167,888 | 26,140,904 | 14,138,587 |
|  | 3 | 4,321,176 | 11,259,105 | 22,162,550 | 15,374,727 | 10,247,773 | 17,817,244 | 24,143,876 | 17,736,556 |
|  | 4 | 2,580,689 | 2,890,398 | 6,882,722 | 13,796,895 | 9,088,456 | 5,822,901 | 10,176,097 | 13,357,787 |
|  | 5 | 1,488,655 | 1,629,131 | 1,631,072 | 3,870,062 | 7,153,127 | 4,428,360 | 2,798,812 | 4,662,082 |
|  | 6 | 1,075,412 | 897,695 | 883,237 | 919,738 | 2,001,306 | 3,506,905 | 2,131,377 | 1,227,227 |
|  | 7 | 988,848 | 592,305 | 426,066 | 446,908 | 410,819 | 854,297 | 1,466,585 | 843,135 |
|  | 8 | 1,382,238 | 544,629 | 281,121 | 215,584 | 199,620 | 175,366 | 357,267 | 580,156 |
|  | 9 | 325,547 | 761,296 | 258,493 | 142,244 | 96,295 | 85,212 | 73,338 | 141,328 |
|  | 10+ | 312,460 | 351,395 | 528,108 | 398,011 | 241,315 | 144,116 | 95,905 | 66,949 |

Table 6.4.5.2.4. Blue whiting. AMCI summary table
SUMMARY TABLE

| Year | Recruits <br> age 1 | SSB | F <br> $(3-7)$ | Catch <br> SOP |
| ---: | ---: | ---: | ---: | ---: |
| 1981 | $3,358,560$ | $2,807,506$ | 0.3087 | 922,980 |
| 1982 | $3,920,661$ | $2,292,762$ | 0.2305 | 550,643 |
| 1983 | $1,450,733$ | $1,845,598$ | 0.2681 | 553,344 |
| 1984 | $19,016,673$ | $1,507,602$ | 0.3431 | 615,569 |
| 1985 | $12,011,290$ | $1,648,462$ | 0.3782 | 678,214 |
| 1986 | $9,429,937$ | $1,887,405$ | 0.5078 | 847,145 |
| 1987 | $9,260,072$ | $1,695,170$ | 0.4585 | 654,718 |
| 1988 | $7,706,865$ | $1,506,629$ | 0.4254 | 552,264 |
| 1989 | $9,015,810$ | $1,447,483$ | 0.5074 | 630,316 |
| 1990 | $21,516,302$ | $1,350,967$ | 0.5093 | 558,128 |
| 1991 | $9,083,595$ | $1,786,877$ | 0.2165 | 364,008 |
| 1992 | $6,104,598$ | $2,403,472$ | 0.2183 | 474,592 |
| 1993 | $5,532,516$ | $2,363,797$ | 0.2025 | 475,198 |
| 1994 | $6,497,651$ | $2,342,592$ | 0.1921 | 457,696 |
| 1995 | $7,573,674$ | $2,176,386$ | 0.2328 | 505,175 |
| 1996 | $20,506,833$ | $2,014,202$ | 0.3056 | 621,104 |
| 1997 | $41,591,472$ | $2,066,203$ | 0.3122 | 639,680 |
| 1998 | $28,588,843$ | $2,845,329$ | 0.4334 | $1,131,954$ |
| 1999 | $19,033,352$ | $3,447,824$ | 0.397 | $1,261,033$ |
| 2000 | $33,972,101$ | $3,492,714$ | 0.4908 | $1,412,449$ |
| 2001 | $47,003,317$ | $3,679,005$ | 0.5399 | $1,771,805$ |
| 2002 | $34,434,779$ | $4,066,588$ | 0.5535 | $1,556,954$ |
| 2003 | $20,000,000$ | $4,294,960$ | 0.6103 | $2,365,319$ |
| 2004 |  | $3,789,678$ |  |  |

Table 6.5.1. Blue Whiting. Input data for the stochastic short-term prediction

## STPR3

Run: AMCI based Short term pred BW 2005
Time and date: May 2
Fbar age range: 3-7
Ochhams razor' stock-recruitment function (p1: 13.8 10*9 individuals; p2: 1.5 million tonnes)

| 2004 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Stock size | Natural mortality | Maturity ogive | Prop. of F bef. spaw. | Prop. of M bef. spaw. | Weight in stock | Exploit. pattern | Weight in catch |
| 1 | from S/R | 0.2 | 0.11 | 0.25 | 0 |  | 0.1468 |  |
| 2 |  | 0.2 | 0.40 | 0.25 | 0 |  | 0.1879 |  |
| 3 |  | 0.2 | 0.82 | 0.25 | 0 |  | 0.392 |  |
| 4 |  | 0.2 | 0.86 | 0.25 | 0 |  | 0.5806 |  |
| 5 |  | 0.2 | 0.91 | 0.25 | 0 |  | 0.6245 |  |
| 6 |  | 0.2 | 0.94 | 0.25 | 0 | sampled | 0.7275 | sampled |
| 7 |  | 0.2 | 1.00 | 0.25 | 0 | from | 0.7275 | from |
| 8 | from AMCI | 0.2 | 1.00 | 0.25 | 0 | table | 0.7275 | table |
| 9 | bootstrap | 0.2 | 1.00 | 0.25 | 0 | 6.3.3.1 | 0.7275 | 6.3.3.1 |
| 10 |  | 0.2 | 1.00 | 0.25 | 0 |  | 0.7275 |  |


| 2004 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Stock <br> size | Natural mortality | Maturity ogive | Prop. of F bef. spaw. | Prop. of M bef. spaw. | Weight in stock | Exploit. pattern | Weight in catch |
| 1 | from S/R | 0.2 | 0.11 | 0.25 | 0 |  | 0.1468 |  |
| 2 | . | 0.2 | 0.40 | 0.25 | 0 |  | 0.1879 |  |
| 3 | . | 0.2 | 0.82 | 0.25 | 0 |  | 0.392 |  |
| 4 | . | 0.2 | 0.86 | 0.25 | 0 |  | 0.5806 |  |
| 5 | . | 0.2 | 0.91 | 0.25 | 0 |  | 0.6245 |  |
| 6 | . | 0.2 | 0.94 | 0.25 | 0 | sampled | 0.7275 | sampled |
| 7 | . | 0.2 | 1.00 | 0.25 | 0 | from | 0.7275 | from |
| 8 | . | 0.2 | 1.00 | 0.25 | 0 | table | 0.7275 | table |
| 9 | . | 0.2 | 1.00 | 0.25 | 0 | 6.3.3.1 | 0.7275 | 6.3.3.1 |
| 10 |  | 0.2 | 1.00 | 0.25 | 0 |  | 0.7275 |  |


| 2005 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Stock <br> size | Natural mortality | Maturity ogive | Prop. of F bef. spaw. | Prop. of M bef. spaw. | Weight in stock | Exploit. pattern | Weight in catch |
| 1 | from S/R | 0.2 | 0.11 | 0.25 | 0 |  | 0.1468 |  |
| 2 | . | 0.2 | 0.40 | 0.25 | 0 |  | 0.1879 |  |
| 3 | . | 0.2 | 0.82 | 0.25 | 0 |  | 0.392 |  |
| 4 | . | 0.2 | 0.86 | 0.25 | 0 |  | 0.5806 |  |
| 5 | . | 0.2 | 0.91 | 0.25 | 0 |  | 0.6245 |  |
| 6 | . | 0.2 | 0.94 | 0.25 | 0 | sampled | 0.7275 | sampled |
| 7 | . | 0.2 | 1.00 | 0.25 | 0 | from | 0.7275 | from |
| 8 | . | 0.2 | 1.00 | 0.25 | 0 | table | 0.7275 | table |
| 9 | . | 0.2 | 1.00 | 0.25 | 0 | 6.3.3.1 | 0.7275 | 6.3.3.1 |
| 10 | . | 0.2 | 1.00 | 0.25 | 0 |  | 0.7275 |  |

Table 6.5.2 Blue Whiting
Probabilistic catch forecast for 2005 assuming status quo F in 2004.
Confidence intervals correspond to 25 and 75 percentiles.
F2004=F2003=0.61(0.56-0.74); median catch 2004=1779 (1661-1915); median SSB 2004=3381 (2927-3851)

|  | F2005 |  |  | SSB 2005 |  | SSB 2006 |  |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Catch 2005 | median | confidence interval | median | confidence interval | median | confidence interval |  |
| 800 | 0.230 | $(0.197-0.269)$ | 3915 | $(3381-4556)$ | 4087 | $(3521-4681)$ |  |
| 900 | 0.262 | $(0.224-0.307)$ | 389 | $(3354-4531)$ | 4003 | $(3431-4589$ |  |
| 1000 | 0.295 | $(0.252-0.346)$ | 3866 | $(3327-4506)$ | 3916 | $(3350-4500)$ |  |
| 1075 | 0.320 | $(0.273-0.376)$ | 3848 | $(3306-4487)$ | 3849 | $(3283-4429)$ |  |
| 1100 | 0.329 | $(0.280-0.387)$ | 3841 | $(3299-4481)$ | 3828 | $(3261-4407)$ |  |
| 1200 | 0.364 | $(0.309-0.428)$ | 3816 | $(3270-4455)$ | 3742 | $(3176-4319($ |  |
| 1300 | 0.400 | $(0.339-0.471)$ | 3789 | $(3241-4429)$ | 3657 | $(3095-4232)$ |  |
| 1400 | 0.436 | $(0.369-0.516)$ | 3761 | $(3213-4400)$ | 3566 | $(3008-4145)$ |  |
| 1500 | 0.474 | $(0.400-0.562)$ | 3733 | $(3185-4272)$ | 3476 | $(2924-4057)$ |  |
| 2000 | 0.681 | $(0.569-0.818)$ | 3587 | $(3019-4223)$ | 3041 | $(2499-3633)$ |  |
| 2500 | 0.929 | $(0.764-1.136)$ | 3416 | $(2845-4058)$ | 2608 | $(2070-3188)$ |  |
| 2930 | 1.183 | $(0.954-1.489)$ | 3252 | $(2664-3903)$ | 2250 | $(1770-2820)$ |  |

Table 6.6.1. Medium term projection for blue whiting spawning stock biomass based on the final AMCI run and 3 different scenarios for mean recruitment and catch regime. SSBs below $\mathrm{B}_{\mathrm{pa}}$ are highlighted.

| Mean recruitment when $\mathrm{SSB}>1.5$ mill. tonnes | Year | Catch 2.0 mill. tonnes |  |  | Catch 1.5 mill. tonnes |  |  | Catch 1.0 mill. tonnes |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 25\% | 50\% | 75\% | 25\% | 50\% | 75\% | 25\% | 50\% | 75\% |
| Fitted 1981-2002 <br> ( 13.8 bill.) | 2005 | 3019 | 3587 | 4223 | 3185 | 3733 | 4373 | 3327 | 3866 | 4506 |
|  | 2010 | 0 | 0 | 0 | 0 | 908 | 1699 | 2465 | 3307 | 4057 |
|  | 2015 | 0 | 0 | 0 | 0 | 0 | 0 | 1515 | 2409 | 3320 |
| Geometric mean 19932002 | 2005 | 3071 | 3641 | 4266 | 3228 | 3792 | 4409 | 3361 | 3921 | 4542 |
|  | 2010 | 0 | 0 | 843 | 1324 | 2354 | 3322 | 3696 | 4691 | 5668 |
| (18.7 bill.)Arithmetic mean 1993-2002(23.3 bill.) | 2015 | 0 | 0 | 0 | 0 | 254 | 2165 | 4001 | 5092 | 6245 |
|  | 2005 | 3111 | 3676 | 4307 | 3261 | 3825 | 4454 | 3395 | 3959 | 4582 |
|  | 2010 | 0 | 1198 | 2399 | 2447 | 3628 | 4797 | 4843 | 6056 | 7205 |
|  | 2015 | 0 | 0 | 0 | 1916 | 3459 | 5038 | 6253 | 7538 | 9028 |

Table 6.9.1 Blue whiting. Total landings, No. of samples, No. of fish measured and No. of fish aged by country and quarter for 2003.

| Country | Quarter | Landings (t) | No. Of samples | No. Fish measured | No. Fish aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 1 | 10088 | 2 | 91 | 91 |
|  | 2 | 21661 | 8 | 401 | 401 |
|  | 3 | 43821 | 9 | 412 | 412 |
|  | 4 | 7365 | 1 | 50 | 50 |
|  | Total | 82935 | 20 | 954 | 954 |
| Faroe Islands | 1 | 42646 | 7 | 1554 | 600 |
|  | 2 | 166556 | 9 | 2047 | 900 |
|  | 3 | 73601 | 7 | 1518 | 700 |
|  | 4 | 47092 | 5 | 1135 | 500 |
|  | Total | 329895 | 28 | 6254 | 2700 |
| France | 1 | 11793 | 0 | 0 | 0 |
|  | 2 |  |  |  |  |
|  | 3 | 1570 | 0 | 0 | 0 |
|  | 4 | 786 | 0 | 0 | 0 |
|  | Total | 14149 | 0 | 0 | 0 |
| Germany | 1 | 3617 | 0 | 0 | 0 |
|  | 2 | 18379 | 6 | 257 | 257 |
|  | 3 | 807 | 0 | 0 | 0 |
|  | 4 |  |  |  |  |
|  | Total | 22803 | 6 | 257 | 257 |
| Iceland | 1 | 7846 | 2 | 200 | 100 |
|  | 2 | 143559 | 46 | 7008 | 3850 |
|  | 3 | 231589 | 69 | 5568 | 3342 |
|  | 4 | 118499 | 29 | 2288 | 1504 |
|  | Total | 501493 | 146 | 15064 | 8796 |
| Ireland | 1 | 22480 | 12 | 2358 | 1296 |
|  | 2 | 28 | 0 | 0 | 0 |
|  | 3 |  |  |  |  |
|  | 4 | 72 | 0 | 0 | 0 |
|  | Total | 22580 | 12 | 2358 | 1296 |
| Norway | 1 | 347863 | 52 | 2488 | 2307 |
|  | 2 | 369542 | 59 | 3926 | 1512 |
|  | 3 | 95729 | 35 | 2746 | 894 |
|  | 4 | 38262 | 20 | 1959 | 257 |
|  | Total | 851396 | 166 | 11119 | 4970 |
| Portugal | 1 | 586 | 85 | 9923 | 203 |
|  | 2 | 734 | 92 | 10669 | 239 |
|  | 3 | 603 | 85 | 9651 | 255 |
|  | 4 | 728 | 80 | 9459 | 208 |
|  | Total | 2651 | 342 | 39702 | 905 |
| Russia | 1 | 63956 | 39 | 6210 | 364 |
|  | 2 | 144372 | 208 | 36810 | 1672 |
|  | 3 | 80700 | 285 | 29016 | 1368 |
|  | 4 | 66291 | 98 | 13626 | 100 |
|  | Total | 355319 | 630 | 85662 | 3504 |
| Scotland | 1 | 20768 | 0 | 0 | 0 |
|  | 2 | 6479 | 0 | 0 | 0 |
|  | 3 | 30 | 0 | 0 | 0 |
|  | 4 | 104 | 0 | 0 | 0 |
|  | Total | 27381 | 0 | 0 | 0 |
| Spain | 1 | 3319 | 30 | 2542 | 300 |
|  | 2 | 3461 | 38 | 2439 | 300 |
|  | 3 | 3379 | 57 | 5680 | 425 |
|  | 4 | 3666 | 66 | 6822 | 425 |
|  | Total | 13825 | 191 | 17483 | 1450 |
| Sweden | 1 | 15 | 0 | 0 | 0 |
|  | 2 | 7627 | 0 | 0 | 0 |
|  | 3 | 57673 | 0 | 0 | 0 |
|  | 4 | 217 | 0 | 0 | 0 |
|  | Total | 65532 | 0 | 0 | 0 |
| The Netherlands | 1 | 22268 | 5 | 1264 | 125 |
|  | 2 | 33967 | 49 | 8349 | 1225 |
|  | 3 | 1021 | 1 | 304 | 25 |
|  | 4 |  | 0 | 0 | 0 |
|  | Total | 57261 | 55 | 9917 | 1375 |
| Grand Total |  | 2347220 | 1596 | 188770 | 26207 |

Table 6.9.2. Blue Whiting. Sampling levels in 2003 per area

| Area | landings | nos samples | nos measured | nos aged |
| :---: | :---: | :---: | :---: | :---: |
| I | 46 | 1 | 49 | 49 |
| IIa | 832,173 | 629 | 74,780 | 8,441 |
| IIb | 16,973 | 0 | 0 | 0 |
| IIIa | 16,162 | 7 | 345 | 345 |
| IVa | 140,992 | 78 | 7,342 | 1,231 |
| IVb | 1,026 | 0 | 0 | 0 |
| IXa | 2,651 | 342 | 39,702 | 905 |
| Va | 82,253 | 29 | 2,966 | 1,722 |
| Vb | 468,269 | 136 | 23,311 | 5,728 |
| VIa | 126,037 | 24 | 2,575 | 1,021 |
| VIb | 337,915 | 75 | 7,544 | 2,452 |
| VIIb | 276 | 0 | 0 | 0 |
| VIIbc | 128,090 | 14 | 750 | 651 |
| VIIc | 40,850 | 10 | 2,081 | 621 |
| VIIg |  | 0 | 0 | 0 |
| VIIgk | 71,262 | 7 | 504 | 200 |
| VIIg-k | 4,945 | 0 | 0 | 0 |
| VIIgk+XII | 19,354 | 36 | 6,217 | 300 |
| VIIh | 142 | 0 | 0 | 0 |
| VIIIabd | 1,380 | 1 | 304 | 25 |
| VIIIc+IXa | 13,825 | 191 | 17,483 | 1,450 |
| VIIj | 2,241 | 2 | 569 | 50 |
| VIIk | 10,624 | 9 | 1,929 | 1,000 |
| XII | 3,854 | 0 | 0 | 0 |
| XIVb | 63 | 5 | 319 | 16 |
| total | 2,321,406 | 1,596 | 188,770 | 26,207 |



Figure 6.2.1. Total catches of blue whiting in 2003 by quarter and ICES rectangle. Grading of the symbols: small dots 10-100 t , white squares 100-1 000 t , grey squares $1000-10000 \mathrm{t}$, and black squares $>10000 \mathrm{t}$.


## Catches in tonnes

-Greater than 10,000 (62)
$\square 1,000$ to 10,000

- 100 to 1,000
- 10 to 100
(150)

Figure 6.2.2. Total catches of blue whiting in 2003 by ICES rectangle. Grading of the symbols: small dots $10-100 \mathrm{t}$, white squares 100-1 000 t , grey squares $1000-10000 \mathrm{t}$, and black squares $>10000 \mathrm{t}$.


Figure 6.4.1.1.1. Density of blue whiting in terms $\mathrm{s}_{\mathrm{A}}$-values $\left(\mathrm{m}^{2} / \mathrm{n}\right.$. mile $\left.^{2}\right)$ based on 5 nm values reported by each of the four research vessels during the international blue whiting spawning stock survey in March-April 2004.


Figure 6.4.1.1.2. Length and age distribution of blue whiting estimated from the international blue whiting spawning stock survey in March-April 2004.


Figure 6.4.1.1.3. Blue whiting biomass in 1000 tonnes, spring 2004. R/V Johan Hjort, March-April 2004.


Figure 6.4.1.1.4. Length and age distribution of blue whiting as estimated by R/V Johan Hjort in March-April 2004.

## Portuguese bottom trawl survey (Summer)



## Spanish Bottom Trawl Surveys



Figure 6.4.2.1 Blue whiting. Mean catch rates in the bottom trawl surveys from the southern area.


Figure 6.4.2.2. Bottom trawl 0-group index (number of blue whiting caught $10^{-3}$ ) from autumn 1996-2003 compared to the following 1 -group spring index lagged one year to match the 0 -group index.

CPUE Spanish pair trawlers


Figure 6.4.3.1 Blue whiting CPUE from Spanish Pair trawlers in ICES Div VIIIc and IXa (North)

CPUE Norway


Figure 6.4.3.2. Blue whiting. Overall aggregated CPUE from the Norwegian directed fisheries in 1982-2003 (tonnes/hour). Vertical bars show strandard deviation and 95\% confidence limits.




Figure 6.4.4.2.1. Blue Whiting. Comparison on blue whiting assessments with the settings from assessment from 2003 and from data available to either WG2003 and 2004 with different versions of AMCI.




Figure 6.4.4.2.2. Blue Whiting. Data exploration with AMCI. Effects of deviations from WG2003 settings including restricting age ranges of survey time series as recommended by SGAMHBW.




Figure 6.4.4.2.3. Blue Whiting. Further data exploration with AMCI , with particular attention on the effects of assumption on year class 1 years (the recruits).
Figure 6.4.4.2.4. Blue Whiting. Diagnostics from AMCI








0 - fit on catch-at-age
1 - fit on Norway spawning acoustic $\mathrm{N}(\mathrm{a}, \mathrm{y})($ 1981-1990)
2 - fit on Norway spawning acoustic $N(a, y($ (1991-2003)
3 - fit on Russian spawning acoustic $N(a, y)(1982-1991)$
4 - fit on Russian spawning acoustic $N(a, y)(1992-2003)$
5 - fit on Norwegian sea acoustic $N(a, y)(1981-1990)$
6 - fit on Norwegian sea acoustic $N(a, y)(1991-2003)$

Figure 6.4.4.3.1
Blue whiting. Profiles of components of the effort-controlled ISVPA loss function.










0 - fit on catch-at-age
1 - fit on Norway spawning acoustic $\mathbf{N}(\mathrm{a}, \mathrm{y})(1981-1990)$
2 - fit on Norway spawning acoustic $N(a, y($ (1991-2003) 3 - fit on Russian spawning acoustic $N(a, y)(1982-1991)$ 4 - fit on Russian spawning acoustic $N(a, y)(1992-2003)$ 5 - fit on Norwegian sea acoustic $\mathrm{N}(\mathrm{a}, \mathrm{y})(1981-1990)$ 6 - fit on Norwegian sea acoustic $N(a, y)(1991-2003)$

Figure 6.4.4.3.2
Blue whiting. Profiles of components of the catch-controlled ISVPA loss function.


Figure 6.4.4.3.3
Blue whiting. Estimates of SSB from effort-controlled and catch-controlled versions of ISVPA.


Blue whiting.
Comparison of reported and theoretical catches (C-controlled ISVPA)

Figure 6.4.4.3.4 Blue whiting. Comparison of theoretical catches with reported values for two versions of ISVPA.


Figure 6.4.4.3.5. Plots of residuals for the effort-controlled ISVPA


Figure 6.4.4.3.6. Plots of residuals for the catch-controlled ISVPA


Blue whiting, f-controlled ISVPA, retrospective runs



Blue whiting, catrch-controlled ISVPA, retrospective runs

Figure 6.4.4.3.7 Comparison of ISVPA retrospective runs for effort-controlled and catch-controlled versions.



Figure 6.4.4.3.8. Results of blue whiting stock assessment by mean of ISVPA




Blue whiting. ISVPA (f-controlled). Boostrap
Figure 6.4.4.3.9. Estimates of uncertainty in the results of blue whiting stock assessment by means of ISVPA.




Figure 6.4.4.3.10 Comparison of results of blue whiting stock assessment by means of ISVPA made in 2003 and 2004.


Figure 6.4.4.4.1. Blue Whiting. SSQ surface plot for final ICA run for blue whiting in 2004.

Separable Model Diagnostics


Figure 6.4.4.4.2. Blue Whiting. Catch residuals and selection pattern for final ICA run for blue whiting in 2004.

| Landings | Fishing Martality |
| :---: | :---: |
| Recruitment | Stack Size |

Figure 6.4.4.4.3. Standard plots from ICA final run (run 22b) for blue whiting in 2004.




Figure 6.4.4.4.4. Uncertainty in the final ICA run (run 26) for blue whiting in 2004. Results of bootstrapping, with $5 \%, 25 \%, 50 \%, 75 \%$ and $95 \%$ percentiles presented.



Figure 6.4.4.4.5 Blue Whiting. Diagnostics from ICA
Figure 6.4.4.4.5 Blu Whitng. Diagnostics fron
O:\Advisory Process\ACFM $\backslash W G R E P S \backslash W G N P B W \backslash R E P O R T S \backslash 2004 \backslash S 6 . D o c$






Figure 6.4.4.5.1 Blue Whiting. Diagnostics from SMS

Figure 6.4.4.5.2 Selection patters for blue whiting catches as estimated by SMS (Each series sums up to 1)


Figure 6.4.4.5.3 Blue whiting, observed yield and predicted yield by SMS model


Figure 6.4.4.5.4 Blue whiting, Uncertainties of estimated stock numbers at age in 2003 and SSB and $F$ as estimated by SMS




Figure 6.4.4.6.1. Blue whiting. Comparison of normalized spawning stock biomass estimates obtained from acoustic surveys and analytical assessments.



$\begin{array}{lllllllllllll}1980 & 1982 & 1984 & 1986 & 1988 & 1990 & 1992 & 1994 & 1996 & 1998 & 2000 & 2002 & 2004\end{array}$

Figure 6.4.5.2.1. Blue Whiting: Summary of the final AMCI assessment.


Figure 6.4.5.2.2. Final AMCI assessment for blue whiting. Relative change in the objective function relative to the minimum for various values of terminal F (year 2003).


Figure 6.4.5.2.3. Final AMCI assessment on blue whiting. Fishing mortality at age and year.

Fishing mortality (5\%, 25\%, 50\%, 75\%, 95\%)


Recruitment, age 1 ( $5 \%, 25 \%, 50 \%, 75 \%$, $95 \%$ )


SSB (5\%, 25\%, 50\%, 75\%, 95\%)


Figure 6.4.5.2.4. An evaluation of uncertainty in AMCI assessment obtained by bootstrapping the residuals of the final AMCI assessment ( 1000 replicates).


Figure 6.4.5.2.5. Interdependence between estimated fishing mortality and spawning stock biomass in the terminal year in AMCI bootstrap replicates. Point estimates for F and SSB obtained through different assessment models.


Figure 6.4.5.2.6. Blue whiting. AMCI, retrospective runs




Figure 6.4.5.2.7. Blue Whiting: Comparisons between final AMCI assessments by the WGNPBW in 2003 and 2004

### 7.1 The fishery

The catches of Icelandic summer-spawning herring from 1983-2003 are given in Tables 7.1.1. No discards were reported for the 2003/2004 season. The main fishery started in September and terminated in January. The catch in September-January was about 120000 tonnes. The catch was taken with traditional purse-seines and pelagic trawls. Only a purse-seine fishery took place off the east -southeast coast of Iceland in September and December but in all the other months both purse-sein and pelagic trawl fishery took place off the east-southeast and west-northwest coast. Because small herring was frequently present in the catches, especially at SE- and SW-Iceland, area closure was common during the 2003/2004 season.

A total of about 46000 tonnes were fished off west and about 80000 tonnes off east Iceland and $61 \%$ of the catch was taken by purse-sein. About $53 \%$ of the catch was reduced to oil and fishmeal. The remainder was either salted or frozen for human consumption.

Until 1990, the herring fishery took place during the last three months of the calendar year, but from 1990-2002 and 2004 the autumn fishery has continued until January or early February of the following year. In 2003 the season was further extended to the end of April and in the summers of 2002 and 2003 an experimental fishery for spawning herring with a catch of about 5000 t each year was conducted at the south coast. In 1994 the fishery started in September. Therefore, all references to the years 1990-1993 imply seasons starting in October of that year, but after that in September. Landings, catches and recommended TACs since 1984 are given in thousand tonnes in Table 7.1.1.

### 7.2 Catch in numbers, weight-at-age and maturity

The catches of the Icelandic summer-spawners in numbers at age for the period 1982-2003 are given in Table 7.2.1. Like last year age is now given as real age instead of rings, as in earlier WG reports.

During the 1995/96-1997/98 seasons, catches were mainly distributed on the 4 year classes from 1988-1991. On the other hand, catches during the 1998/99 and 1999/2000 seasons were dominated by the strong 1994 year class. In 2000/2001 the very strong 1994 and 1996 year classes were most abundant in the catch, but in 2001/2002 the 1996 year class was the most abundant. In 2002/2003 the 1998 and 1999 year classes were the most common in the catch followed by the 1996 and 1994 year classes. In 2003/2004 the very strong year classes of 1999 and 2000 dominated in the catches with $42 \%$ and $35 \%$ in numbers respectively (Figure 7.2.1). The 1998 year class was $7.5 \%$ in numbers in the catch but the very strong year classes of 1994 and 1996 which dominated in the catches in 1998/1999-2001/2002 were only $1.4 \%$ and $3.6 \%$ in numbers respectively.

The weight at age for each year is given in Table 7.2.2. The weights from the catch are used as the weights for the stock. Proportion mature at age is given in Table 7.2.3.

### 7.3 Acoustic surveys

The Icelandic summer-spawning herring stock has been monitored by annual acoustic surveys since 1974 (Table 7.3.1). These surveys have been conducted in October-December or January. A survey was conducted in late November 2003 at the east coast of Iceland. On the traditional fishing grounds off the east coast the survey recorded about 200000 t of which 177000 t were adult herring. Another survey to the west of Iceland was conducted in late January 2004. A total of 546000 tonnes were recorded of which adult herring was estimated to be 431000 t . The total estimation of the adult stock was therefore 608000 t .

The 2003/2004 acoustic assessment surveys confirmed that the 1999 year class is strong (Table 7.3.1) and it seems that the 2000 year class is also strong (Figure 7.3.1).

The sum of results obtained in winter 2003/2004 acoustic surveys have been used as basis for the present assessment (Table 7.3.1).

Jakobsson et al. (1993) formally tested whether it was feasible to maintain a one-to-one relationship between acoustic and VPA estimates of stock size. It was found that a modification of the target strength, from TS=21.7 log (L) - 75.5 dB to $\mathrm{TS}=20 \log (\mathrm{~L})-72 \mathrm{~dB}$, gave a much better fit between the two data sets. The latter target strength has been the basis of calculations of stock abundance from acoustic survey data since 1993.

This year an update assessment should be performed for the Icelandic summer spawning herring. The assessment was made in AMCI 2.2 as in last year and compared to last years assessment (Figure 7.4.1) As can been seen the new assessment fits well regarding the fishing mortality, but that the spawning stock biomass has been overestimated in last year's assessment. This fits quite well, as the larger herring was not found in this year acoustic survey. The results from the update assessment are shown in table 7.4.1. The update assessment was also run in AMCI 2.3a. From figure 7.4.1 it can be seen that the spawning stock biomass is constantly below the estimates derived with AMCI 2.2. This is due to a bug in AMCI 2.2. Last year an error was made in estimating selection for the catch and the catchability in the survey. Selection above age 5 was supposed to be flat, but in reality it was estimated for all age groups. The same is true for the catchability. The model was over parameterized and results should therefore be interpretated with caution.

### 7.5 Data exploration

In previous years when the ADAPT-type of VPA (which is different from NFT ADAPT 2.1.7 used later here) was used there was a general trend to overestimate SSB and underestimate F. The terminal F gain by that method was not in line with the mortality sign derived from the catch data. In last year (2003) an AMCI assessment was made and seemed to produce a more reliable F. But at the last minute it was discovered that the residuals derived from the AMCI model showed a cyclic pattern, which could possibly mean some time trend in the model. The assessment was accepted but it was decided to look closer into this pattern at later time. Before assessment runs would be done this year (2004) it was decided to look first at the input data.

### 7.5.1 Analysis of catch-at-age in commercial data and in the acoustic survey

Using catchdata from 1981 catchcurves were plotted (Figures 7.5.1.1-7.5.1.2). From them it can be seen that the total mortality sign is between 0.3 and 0.4 provided that the effort has been the same the whole time. It can also be seen, that in later years the fish is fully recruited to the fishery at younger ages. By using data from the years 1981-2003 and ages 2-15 a Shepherd-Nicholson model was run for the catchdata resulting in $\mathrm{CV}=0.82$ which is high in catch in numbers, but can result from variable selection. It was discovered that agegroup 15 was sometimes used as a plus group and sometimes not, so it was excluded from further dataexploration. The CV was estimated as $0.40,0.45,0.53$ and 0.48 when using ages 4-11, 4-12, 3-12 and 3-11 respectively.

Using acoustic survey data from 1981-2004 catchcurves were also plotted (Figures 7.5.1.3-7.5.1.4). They are a bit difficult to interpret. The lines are zigzagged which can be a result of a faulty agereadings, but the mortality sign could be somewhere around 0.4 provided that the catchability at age is the same. However in the last year (2004) there is a huge drop in the curves for older year classes, as the survey did not find the oldest part of the stock. Until now in 2004 only ages 6 and older have been used in the tuning in the assessments, whereas until 2002 the sum of 6 years and older was used. A comparison between ages in the survey was made by fitting a line through the origin (Figure 7.5.1.5) and the correlation from the regression tabulated (Table 7.5.1.1). Based on Table 7.5.1.1 ages 3-9 could be used as age disaggregated indices in an assessment. Ages 10 and 11 could possibly be used, as the correlation is moderate for these ages. A biomass index was made for the acoustic survey. The weights from the catchdata are used as survey weights. The weights were summed over ages 3-9 and 3-11. The index is almost the same for both agegroup categories (Figure 7.5.1.6).

### 7.5.2 Data exploration with NFT ADAPT 2.1.7.

Input data:

- $\quad \mathrm{c} @ a$, years 1981-2003, age groups 2-12. Informations from age groups 13 and above ignored.
- U@a, years 1981-2004, age groups 3-9.


## Excercises performed:

- Run 1: Tuning range 1981-2004, age groups 3-9
- Run 2: Tuning range 1988-2004, age groups 3-9
- Run 3: Tuning range 1988-2004, age groups 3 and 4


## Results:

Run 1 and 2 indicate that there is a discrepancy in the two data sets with retrospective estimates of catchability showing an increasing trend (Figure 7.5.2.1). Given the model assumption of constant catchability this results in a persistent overestimation of biomass and underestimation of fishing mortality.

From run 2 there were indications that the catchability trend may only be in age groups 5-9 but that trends in age groups 3 and 4 may be less pronounced or non existent. When tuning with only age groups 3 and 4 (run 3 ) the retrospective pattern in biomass and fishing mortality is considered resonable in the sense that the pattern is not just one sided (Figures 7.5.2.2 and 7.5.2.3). The biomass of age groups 4 and above in 2003 is estimated to be 732 kt with at standard deviation of $258 \mathrm{kt}(\mathrm{CV}=0.35$, estimates based on 1000 bootstrap runs).

Although change in hidden mortality may a priori be an equally likely hypothesis as change in catchability it cannot be explored with the tool used. If change in mortality is suspected it may only be applicable to older fish. It is considered less likely that the catchability has increased as less survey effort has been a restricting factor since the herring stock has become more widely distributed.

### 7.5.3 Data exploration with TSA

## Input data:

- c@a, years 1986-2003, age groups 4-12. Informations from age groups 13 and above ignored as well as age groups 2 and 3.
- U@a, years 1988-2004, age groups 4-8, which are shifted towards the end of the year before, so they are then ages 3-7 at the end of the years 1987-2003.


## Excercises performed:

- Run 1: Only catch data used
- Run 2: Catch- and surveydata used. Catchability in the survey estimated by using random walk.
- Run 3: Catch- and surveydata used. No random walk term.
- Run 4: Natural mortality is estimated. A model is run where $M$ is fixed for all age groups but changes in time according to the model $\mathrm{M}(\mathrm{t})=0.1+\Theta \mathrm{g}(\mathrm{t})$, where $\mathrm{g}(\mathrm{t})=0$ for $\mathrm{t}=1986$-1993, $\mathrm{g}(\mathrm{t})=(0.2,0.5,0.8)$ in 1994-1996 and $\mathrm{g}(\mathrm{t})=1$ from 1997-2003. The parameter $\Theta$ is estimated.


## Results:

Run 2 is significantly better than Run 3 and is regarded as the best run of the first three runs. The biomass of age groups 4 to 12 in 2003 is estimated to be 555 kt with a standard deviation of 35.8 kt . The unweighted F is estimated to be 0.135 for age groups $5-12$ with a standard deviations of $\log (\mathrm{F})$ as 0.094 .

The results for the log likelihood function from run 4 is as follow:

|  | M is fixed | M estimated |
| :--- | :--- | :--- |
| catchability fixed | 0.00 | 5.28 |
| catchablility estimated | 3.34 | 6.09 |

One parameter is significant at approximately $\mathrm{P}=0.05$ when the likelihood function increases by 2.0 and an increase of 3.0 is needed so that addition of 2 parameters is significant. According to these values a model where M is fixed and catchability is fixed is rejected. More is gained by estimating variable M by time than by estimating catchability. A model where both M and catchability is estimated is not significant better than model where only variable M is estimated. So this means that there is no reason to reject that the catchability is unchanged. When using this model with variable M by years the biomass of agegroups 4 to 12 in 2003 is 583.5 kt with a standard deviation of 39.7 kt . The
unweighted F for age groups $5-12$ in 2003 is estimated as 0.171 with the standard deviation of the $\log (\mathrm{F})$ as 0.108 . (Estimation of standard deviations of F and stock values in TSA only represent errors from the random variations but assume that the true model is applied. The actual uncertainty is therefore considerably higher)

Althoug the model behaves better with variable M by years it isn't thereby proven that a change in the natural mortality is real. On the other hand it is not considered unlikely as the herring has changed its distribution from being in Fjords and close to the coast to being almost all around the country in deeper waters since 1997/1998.

### 7.5.4 Data exploration with AMCI 2.3a

Input data:

- $\quad$ @a, years 1981-2003, age groups 2-12+. Last year (2003) it was recommended in the technical minutes to try to use a plus group to diminish for effects from misreading of scales as it is very difficult to read scales from old herring, but until now ages 2-15 has been used.
- U@a, years 1981-2004, age groups 3-9


## Excercises performed:

In all runs the objective function was: $\sum \lambda_{a}\left(\log \left(c_{a}\right)-\log \left(\hat{c}_{a}\right)\right)+\sum(\log (y)-\log (\hat{y}))+\sum\left(\log \left(s_{a}\right)-\log \left(\hat{s}_{a}\right)\right)$ where $c_{a}$ and $\hat{c}_{a}$ are the observed versus the fitted catch numbers by age, y and $\hat{y}$ are observed and fitted yield and $s_{a}$ and $\hat{s}_{a}$ are the observed versus the fitted survey values by age. $\lambda_{\mathrm{a}}=0.5$ for age 2 , otherwise set to 1 .

- Run 1: Uses all data. Selection was estimated parametric by age with the condition of flat selection for ages 5 and older. Selection was allowed to change with time for youngest ages. The gainfactor was set to 0.2 for ages 2 and 3 and 0.1 for ages 4 and 5 . Catchablility was estimated parametric at age for ages 3, 4 and 5 and kept flat above that and kept the same for all years.
- Run 2: Catchdata used from years 1986-2003. Tuning range 1988-2004. Selection was estimated with the same conditions as in run 1 , but the gainfactor was set to $0.4,0.3,0.2$ and 0.1 for ages $2,3,4$ and 5 respectively. Catchability was estimated like in run 1.
- Run 3: Catchdata used from years 1986-2003. Tuning range 1988-2004. Three seperable periods were used, 1986-1991, 1992-1997 and 1998-2003. Gain factor was 0 . Catchability was estimated like in run 1.
- Run 4: Catchdata used from years 1986-2003. Tuning range 1988-2004. The same seperable periods were used as in run 3, but in the middle period the selection was estimated parametric at age, except for the plus group that was set to the same selection as age 11 , and then allowed to change gradually by setting the gainfactors as $0.4,0.3,0.2$ and 0.1 for ages $2,3,4$ and 5 and older. Catchablity was estimated like in run 1.


## Results:

In all the runs recruitment at age 2 in 2003 could not be estimated reliably, so it was set to a geometric mean of 650 derived in last year. It was observed that the geometric mean value for the new assessments is 520, but it doesn't have any effect on the assessments. In run 2 selection at age 2 in the first year was poorly estimated and in run 3 and 4 there was a year trend in the survey residuals. So run 1 is regarded as the best amci run. As seen from the diagnostic plot for run 1 (Figure 7.5.4.1) it can be seen that the fitted catch doesn't fit the observed very well around 1990 and 2000. There can be a plausible explanation for the discrepancy around 1990. There was the change in the assessment from the fishing year to a fishing season, so there are most likely data discrepancies in the years 1990 and 1991. The run 1 doesn't either follow the survey very well in the last years. Recruitment is estimated record high for the year classes 1999 and 2000. A retrospective run was made for this assessment (Figure 7.5.4.2)

### 7.5.5 Adapt in AD-Model builder

Input data:

- $\quad$ @a, years 1981-2003, age groups 2-12.
- U@a, years 1988-2004, age groups 3-12

A cohort model using Popes approximation was used. Catchability was fixed from age 8 onwards. CV in the survey data was estimated for each age agegroup.

## Results:

The results from this run are shown in figure 7.5.5.1. The recruitment is estimated very high for the 1999 and the 2000 year classes. It doesn't follow the survey very well and in the residual plot for the survey it can be seen that the model expected higher numbers for the older age groups in the last survey. A retrospective run was made for this run (Figure 7.5.5.2). As can been seen from it there has been a tendency to overestimate the spawning stock in the past years.

### 7.5.6 Comparisons between models

The assessments made with the different models were compared (Figure 7.5.6.1) as well as with the assessment made last year (2003) run in AMCI 2.3, although it is not quite comparable with the other exploratory assessment. Last year age groups 2 to 15 were used in the tuning fleet but now only ages 3 to 9 have been used. And now terminal fishing mortality is computed for ages $5-10$ instead of $5-15$, as 12 is used as a plus group in this year's assessment. In general all the assessments have the same trend. The F values from last year's assessment are somewhat different from the Fs of the other 4 assessments, and the F-values for the NFT Adapt run are different in the first years. For the Adapt- and the TSA assessment biomass of $4+$ is shown. Usually about $90 \%$ of age group 4 is mature, so it can easily be compared to the spawning stock. Regarding the spawning stock all runs show similar trend, except that TSA is a level higher in the last year and NFT Adapt is level higher in earlier years. The recruitment is very similar over the years and all runs estimate high recruitment for the 1999 and 2000 year classes. In this TSA run a time variable natural mortality is used and the update assessment is overparametrized so they have to be left out. As the other models tested are showing the same trend both in SSB and Fs it was not considered justifiable to replace the assessment model from last year.

### 7.6 Final assessment

Run 1 from the AMCI was chosen as the final assessment. According to this assessment the spawning stock biomass was about 490 thousand tonnes in 2003 (Table 7.6.1). The annual unweighted fishing mortality was 0.25 . Stock numbers and fishing mortalities are shown in Tables 7.6.2 and 7.6.3.

### 7.7 Short term projections

AMCI is a forward running program and if the AMCI is used for short-term projection using the stock numbers from the AMCI run a TAC of 160000 t would be obtained. Instead a deterministic short-term projection was made using the MFDP program. The stock numbers for all ages from the AMCI were used except the estimate of the 4 year old in 2004 (the 2000 year class) which is considered unrealistically high. This year class was mixed heavily with the 1999 year class on the fishing grounds. Because of relatively slow growth of the 1999 year class reference length for area closures was lowered compared to previous years and this may have resulted in too high a signal of this year class in the fishery. Although there is indication from surveys and catch data that the 2000 year class is large this indication does not point to a historically large year class as estimated in the assessment run. Furthermore it has been shown that there is a tendency to overestimate the stock biomass when large year classes are entering the stock (Figure 7.5.5.2). It is estimated in all runs higher than the 1999 year classes. The size of the 2000 year class was, therefore, set at the same level as the 1999 year class in the year 2004. Selection is the mean selection in the last 3 years. Proportion of M before spawning is 0.5 and proportion of F before spawning is 0 . The input data is shown in table 7.7.1 and the results from the run in table 7.7.2 Fishing at $0.22\left(=\mathbf{F}_{0.1}\right)$ would correspond to a catch of 106 thousand tonnes in 2004/2005 season.

### 7.7.1 Comments to the assessment

It has been pointed out that there are discrepancies in the catch- and survey datasets, which make it difficult to assess the stock. A project is ongoing, where all scales are being reread back to 1992.

### 7.8 Management consideration

During the last 20 years the Icelandic summer-spawning herring stock has been managed at levels corresponding fairly closely to fishing at $\mathbf{F}_{0.1}$.

The Working Group points out that managing this stock at an exploitation rate at or near $\mathbf{F}_{0.1}$ has been successful in the past. Thus the Working Group agreed in 1998 with the SGPAFM on using $\mathbf{F}_{\mathrm{pa}}=\mathbf{F}_{0.1}=0.22, \mathbf{B}_{\mathrm{pa}}=\mathbf{B}_{\mathrm{lim}} * \mathrm{e}^{1.645 \sigma}=300000 \mathrm{t}$ where $\mathbf{B}_{\mathrm{lim}}=200000 \mathrm{t}$. The Study Group on Precautionary Reference Points for Advice on Fishery Management met in February this year and concluded that it was not considered relevant to change the $\mathbf{B}_{\mathrm{lim}}$ from 200000 t . The present working group agrees with this conclusion.

## $7.10 \quad$ Sampling

| Investigation | No. of samples | Length measured individuals | Aged individuals |
| :--- | :--- | :--- | :--- |
| Fishery | 42 | 2402 | 2223 |
| Acoustic, wintering area | 22 | 6408 | 3395 |

Table 7.1.1 Icelandic summer spawners. Landings, catches and recommended TACs in thousand tonnes.

| Year | Landings | Catches | Recommended <br> TACs |
| :--- | ---: | ---: | ---: |
| 1984 | 50.3 | 50.3 | 50.0 |
| 1985 | 49.1 | 49.1 | 50.0 |
| 1986 | 65.5 | 65.5 | 65.0 |
| 1987 | 73.0 | 73.0 | 70.0 |
| 1988 | 92.8 | 92.8 | 100.0 |
| 1989 | 97.3 | 101.0 | 90.0 |
| $1990 / 1991$ | 101.6 | 105.1 | 90.0 |
| $1991 / 1992$ | 98.5 | 109.5 | 79.0 |
| $1992 / 1993$ | 106.7 | 108.5 | 86.0 |
| $1993 / 1994$ | 101.5 | 102.7 | 90.0 |
| $1994 / 1995$ | 132.0 | 134.0 | 120.0 |
| $1995 / 1996$ | 125.0 | 125.9 | 110.0 |
| $1996 / 1997$ | 95.9 | 95.9 | 100.0 |
| $1997 / 1998$ | 64.7 | 64.7 | 100.0 |
| $1998 / 1999$ | 87.0 | 87.0 | 90.0 |
| $1999 / 2000$ | 92.9 | 92.9 | 100.0 |
| $2000 / 2001$ | 100.3 | 100.3 | 110.0 |
| $2001 / 2002$ | 95.3 | 95.3 | 125.0 |
| $2002 / 2003$ | 92.7 | 93.6 | 105.0 |
| $2003 / 2004 *$ | 125.2 | 125.2 | 110.0 |

*Preliminary

Table 7.2.1 Icelandic summer spawners. Catch in numbers (millions) and total catch in weight (thous. tonnes). 1982 refers to season 1982/1983.

| Age/Year | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.454 | 1.475 | 0.421 | 0.112 | 0.100 | 0.029 | 0.879 |
| 3 | 19.187 | 22.499 | 18.015 | 12.872 | 8.172 | 3.144 | 4.757 |
| 4 | 28.109 | 151.718 | 32.244 | 24.659 | 33.938 | 44.590 | 41.331 |
| 5 | 38.280 | 30.285 | 141.354 | 21.656 | 23.452 | 60.285 | 99.366 |
| 6 | 16.623 | 21.599 | 17.043 | 85.210 | 20.681 | 20.622 | 69.331 |
| 7 | 38.308 | 8.667 | 7.113 | 11.903 | 77.629 | 19.751 | 22.955 |
| 8 | 43.770 | 14.065 | 3.916 | 5.740 | 18.252 | 46.240 | 20.131 |
| 9 | 6.813 | 13.713 | 4.113 | 2.336 | 10.986 | 15.232 | 32.201 |
| 10 | 6.633 | 3.728 | 4.517 | 4.363 | 8.594 | 13.963 | 12.349 |
| 11 | 10.457 | 2.381 | 1.828 | 4.053 | 9.675 | 10.179 | 10.250 |
| 12 | 2.354 | 3.436 | 0.202 | 2.773 | 7.183 | 13.216 | 7.378 |
| 13 | 0.594 | 0.554 | 0.255 | 0.975 | 3.682 | 6.224 | 7.284 |
| 14 | 0.075 | 0.100 | 0.260 | 0.480 | 2.918 | 4.723 | 4.807 |
| 15 | 0.211 | 0.003 | 0.003 | 0.581 | 1.788 | 2.280 | 1.957 |
| Catch | 56.528 | 58.867 | 50.304 | 49.368 | 65.500 | 75.439 | 92.828 |


| Age/year | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 3.974 | 11.009 | 35.869 | 12.006 | 0.869 | 6.225 | 7.411 |
| 3 | 22.628 | 14.345 | 92.758 | 79.782 | 35.560 | 110.079 | 26.221 |
| 4 | 26.649 | 57.024 | 51.047 | 131.543 | 170.106 | 99.377 | 159.170 |
| 5 | 77.824 | 34.347 | 87.606 | 43.787 | 87.363 | 150.310 | 86.940 |
| 6 | 188.654 | 77.819 | 33.436 | 56.083 | 25.146 | 90.824 | 105.542 |
| 7 | 43.114 | 152.236 | 54.840 | 41.932 | 28.802 | 23.926 | 74.326 |
| 8 | 8.116 | 32.265 | 109.418 | 36.224 | 18.306 | 20.809 | 20.076 |
| 9 | 5.897 | 8.713 | 9.251 | 44.765 | 24.268 | 19.164 | 13.797 |
| 10 | 7.292 | 4.432 | 3.796 | 9.244 | 14.318 | 17.973 | 8.873 |
| 11 | 4.780 | 4.287 | 2.634 | 2.259 | 3.639 | 16.222 | 9.140 |
| 12 | 3.449 | 2.517 | 1.826 | 0.582 | 0.878 | 2.955 | 7.079 |
| 13 | 1.410 | 1.226 | 0.516 | 0.305 | 0.300 | 1.433 | 2.376 |
| 14 | 0.844 | 1.019 | 0.262 | 0.203 | 0.200 | 0.345 | 0.927 |
| 15 | 0.348 | 0.610 | 0.298 | 0.102 | 0.100 | 0.345 | 0.124 |
| Catch | 101.000 | 105.097 | 109.489 | 108.504 | 102.741 | 134.003 | 125.851 |


| Age/Year | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 1.100 | 9.323 | 16.161 | 0.629 | 7.958 | 10.206 | 14.149 | 23.470 |
| 3 | 18.723 | 27.072 | 37.787 | 43.537 | 52.921 | 23.944 | 70.982 | 235.988 |
| 4 | 45.304 | 28.397 | 151.853 | 65.871 | 131.153 | 76.666 | 78.395 | 283.519 |
| 5 | 92.948 | 29.451 | 42.833 | 145.127 | 44.334 | 107.849 | 43.905 | 50.527 |
| 6 | 69.878 | 42.267 | 19.872 | 24.653 | 102.925 | 46.646 | 57.266 | 25.898 |
| 7 | 86.261 | 35.285 | 30.280 | 20.614 | 10.962 | 51.585 | 21.433 | 24.010 |
| 8 | 37.447 | 28.506 | 22.572 | 25.853 | 9.312 | 18.504 | 42.272 | 11.330 |
| 9 | 13.207 | 21.828 | 32.779 | 21.163 | 17.218 | 11.356 | 9.668 | 9.743 |
| 10 | 6.854 | 8.160 | 14.366 | 14.436 | 9.471 | 7.933 | 4.632 | 1.429 |
| 11 | 4.012 | 3.815 | 4.802 | 6.973 | 7.610 | 8.547 | 6.429 | 3.750 |
| 12 | 1.672 | 1.696 | 2.199 | 2.164 | 1.930 | 5.090 | 7.839 | 3.479 |
| 13 | 4.179 | 6.570 | 1.084 | 2.426 | 5.199 | 4.346 | 9.738 | 1.743 |
| 14 | 1.672 | 1.378 | 5.081 | 0.473 | 0.552 | 1.611 | 4.478 | 1.272 |
| 15 | 0.100 | 1.802 | 3.036 | 0.961 | 0.166 | 0.864 | 4.537 | 0.816 |
| Catch | 95.882 | 64.682 | 86.998 | 92.896 | 100.332 | 95.278 | 93.601 | 125.2 |

Table 7.2.2 Icelandic summer spawners. Weight at age (g).

| Age/Year | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 59 | 49 | 53 | 60 | 60 | 75 | 63 |  |
| 3 | 132 | 131 | 146 | 140 | 168 | 157 | 130 |  |
| 4 | 180 | 189 | 219 | 200 | 200 | 221 | 206 |  |
| 5 | 218 | 217 | 266 | 252 | 240 | 239 | 246 |  |
| 6 | 260 | 245 | 285 | 282 | 278 | 271 | 261 |  |
| 7 | 309 | 277 | 315 | 298 | 304 | 298 | 290 |  |
| 8 | 329 | 315 | 335 | 320 | 325 | 319 | 331 |  |
| 9 | 356 | 322 | 365 | 334 | 339 | 334 | 338 |  |
| 10 | 370 | 351 | 388 | 373 | 356 | 354 | 352 |  |
| 11 | 407 | 334 | 400 | 380 | 378 | 352 | 369 |  |
| 12 | 437 | 362 | 453 | 394 | 400 | 371 | 389 |  |
| 13 | 459 | 446 | 469 | 408 | 404 | 390 | 380 |  |
| 14 | 430 | 417 | 433 | 405 | 424 | 408 | 434 |  |
| 15 | 472 | 392 | 447 | 439 | 430 | 437 | 409 |  |


| Age/Year | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 63 | 75 | 74 | 63 | 74 | 67 | 69 | 78 |
| 144 | 119 | 139 | 144 | 150 | 135 | 129 | 140 |
| 190 | 198 | 188 | 190 | 212 | 204 | 178 | 166 |
| 232 | 244 | 228 | 232 | 245 | 249 | 236 | 208 |
| 276 | 273 | 267 | 276 | 288 | 269 | 276 | 258 |
| 317 | 286 | 292 | 317 | 330 | 302 | 292 | 294 |
| 334 | 309 | 303 | 334 | 358 | 336 | 314 | 312 |
| 346 | 329 | 325 | 346 | 373 | 368 | 349 | 324 |
| 364 | 351 | 343 | 364 | 387 | 379 | 374 | 360 |
| 392 | 369 | 348 | 392 | 401 | 398 | 381 | 349 |
| 444 | 387 | 369 | 444 | 425 | 387 | 400 | 388 |
| 399 | 422 | 388 | 399 | 387 | 421 | 409 | 403 |
| 419 | 408 | 404 | 419 | 414 | 402 | 438 | 385 |
| 428 | 436 | 396 | 428 | 420 | 390 | 469 | 420 |


| Age/Year | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | $2004^{*}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 62 | 78 | 64 | 58 | 78 | 80 | 66 | 66 |
| 3 | 137 | 147 | 143 | 158 | 140 | 149 | 149 | 149 |
| 4 | 197 | 184 | 211 | 214 | 217 | 202 | 183 | 183 |
| 5 | 234 | 213 | 236 | 256 | 242 | 245 | 225 | 225 |
| 6 | 270 | 246 | 268 | 284 | 281 | 275 | 255 | 255 |
| 7 | 299 | 286 | 300 | 326 | 294 | 311 | 283 | 283 |
| 8 | 323 | 314 | 318 | 333 | 309 | 325 | 328 | 328 |
| 9 | 342 | 341 | 349 | 366 | 339 | 347 | 347 | 347 |
| 10 | 358 | 351 | 347 | 383 | 350 | 383 | 328 | 328 |
| 11 | 363 | 354 | 377 | 402 | 367 | 390 | 368 | 368 |
| 12 | 373 | 350 | 359 | 405 | 375 | 402 | 387 | 387 |
| 13 | 412 | 372 | 403 | 422 | 403 | 442 | 385 | 385 |
| 14 | 394 | 400 | 408 | 406 | 426 | 463 | 380 | 380 |
| 15 | 429 | 437 | 445 | 444 | 425 | 453 | 430 | 430 |

[^5]Table 7.2.3 Icelandic summer spawners. Proportion mature at age.

| Age/Year | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 | 0.000 | 0.010 | 0.000 | 0.030 | 0.010 | 0.045 | 0.060 |
| 4 | 0.640 | 0.820 | 0.900 | 0.890 | 0.870 | 0.900 | 0.930 |
| 5 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 6 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 7 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 8 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 9 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 10 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 11 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 12 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 13 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 14 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 15 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |


| Age/Year | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 | 0.000 | 0.013 | 0.020 | 0.049 | 0.054 | 0.157 | 0.049 |
| 4 | 0.780 | 0.720 | 0.930 | 0.999 | 1.000 | 0.982 | 0.990 |
| 5 | 1.000 | 1.000 | 1.000 | 1.000 | 0.992 | 0.998 | 1.000 |
| 6 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 7 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 8 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 9 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 10 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 11 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 12 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 13 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 14 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 15 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |


| Age/Year | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | $2004^{*}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 | 0.160 | 0.265 | 0.074 | 0.279 | 0.101 | 0.190 | 0.286 | 0.192 |
| 4 | 0.925 | 0.935 | 0.879 | 0.831 | 0.981 | 0.734 | 0.795 | 0.837 |
| 5 | 0.989 | 0.995 | 0.977 | 0.992 | 0.997 | 0.898 | 0.962 | 0.952 |
| 6 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 7 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 8 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 9 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 10 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 11 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 12 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 13 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 14 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |
| 15 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 | 1.000 |

* Predicted (mean of 2001-2003)

Table 7.3.1 Acoustic estimates (in millions) of the Icelandic summer spawning herring, 1974-2003. The surveys are conducted in October-December or January. The year given is the following year, i.e. if the survey is conducted in the season 1973/1974, then 1974 is given.

|  | Ages |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
| 1974 | -1 | 154 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1975 | -1 | 5 | 137 | 19 | 21 | 2 | 2 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1976 | -1 | 136 | 20 | 133 | 17 | 10 | 3 | 3 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1977 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1978 | -1 | 212 | 424 | 46 | 19 | 139 | 18 | 18 | 10 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1979 | -1 | 158 | 334 | 215 | 49 | 20 | 111 | 30 | 30 | 20 | -1 | -1 | -1 | -1 | -1 |
| 1980 | -1 | 19 | 177 | 360 | 253 | 51 | 41 | 93 | 10 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1981 | 625 | 361 | 462 | 85 | 170 | 182 | 33 | 29 | 58 | 10 | -1 | -1 | -1 | -1 | -1 |
| 1982 | -1 | 17 | 75 | 159 | 42 | 123 | 162 | 24 | 8 | 46 | 10 | -1 | -1 | -1 | -1 |
| 1983 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1984 | -1 | 171 | 310 | 724 | 80 | 39 | 15 | 27 | 26 | 10 | 5 | 12 | -1 | -1 | -1 |
| 1985 | -1 | 28 | 67 | 56 | 360 | 65 | 32 | 16 | 17 | 18 | 9 | 7 | 4 | 5 | 5 |
| 1986 | 201 | 652 | 208 | 110 | 86 | 425 | 67 | 41 | 17 | 27 | 26 | 16 | 6 | 6 | 1 |
| 1987 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1988 | 406 | 126 | 352 | 836 | 287 | 53 | 37 | 76 | 25 | 21 | 14 | 17 | 8 | 6 | 3 |
| 1989 | 370 | 725 | 181 | 249 | 381 | 171 | 42 | 23 | 30 | 16 | 10 | 9 | 5 | 3 | 2 |
| 1990 | -1 | 178 | 593 | 177 | 302 | 538 | 185 | -1 | -1 | -1 | 18 | -1 | -1 | -1 | -1 |
| 1991 | 710 | 805 | 227 | 304 | 137 | 176 | 387 | 40 | 10 | 2 | -1 | -1 | -1 | -1 | -1 |
| 1992 | 465 | 745 | 850 | 353 | 273 | 94 | 81 | 210 | 32 | 11 | -1 | 17 | -1 | -1 | -1 |
| 1993 | 1418 | 254 | 858 | 687 | 160 | 99 | 87 | 44 | 92 | 39 | -1 | -1 | -1 | -1 | -1 |
| 1994 | 183 | 234 | 533 | 860 | 443 | 55 | 69 | 43 | 86 | 55 | 2 | -1 | 6 | -1 | -1 |
| 1995 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1996 | 845 | 98 | 165 | 515 | 316 | 361 | 166 | 110 | 52 | 29 | 16 | 27 | 19 | 8 | 2 |
| 1997 | 266 | 792 | 65 | 139 | 459 | 280 | 410 | 150 | 101 | 50 | 35 | 15 | 65 | 32 | -1 |
| 1998 | 1629 | 237 | 716 | 100 | 116 | 240 | 161 | 130 | 97 | 35 | 15 | 11 | 43 | 8 | 15 |
| 1999 | -1 | -1 | 188 | 790 | 240 | 101 | 73 | 47 | 77 | 47 | 10 | 10 | -1 | 22 | -1 |
| 2000 | 1069 | 527 | 740 | 296 | 606 | 99 | 71 | 164 | 108 | 98 | 15 | 44 | 5 | 13 | 7 |
| 2001 | 2832 | 101 | 561 | 1069 | 323 | 609 | 30 | 31 | 38 | 13 | 18 | 6 | 9 | 4 | 1 |
| 2002 | 561 | 942 | 247 | 187 | 265 | 173 | 302 | 69 | 48 | 55 | 54 | 16 | 18 | 1 | -1 |
| 2003 | -1 | 1642 | 1800 | 549 | 221 | 257 | 159 | 273 | 97 | 44 | 43 | 13 | 32 | 2 | -1 |
| 2004 | -1 | 949 | 1529 | 1941 | 316 | 108 | 90 | 34 | 47 | 4 | 3 | 5 | 8 | 3 | 2 |

Table 7.4.1 Results from an update assessment in 2004. Run id 20040502 173313.344

## SUMMARY TABLE

| Year | Recruits <br> age 2 | SSB | $F$ <br> $5-15$ | Catch <br> SOP |
| ---: | ---: | ---: | ---: | ---: |
| 1981 | 817853 | 246943 | 0.1273 | 39461 |
| 1982 | 290683 | 253260 | 0.1957 | 56472 |
| 1983 | 301676 | 274056 | 0.1390 | 58694 |
| 1984 | 532266 | 274503 | 0.0978 | 50132 |
| 1985 | 1283269 | 304134 | 0.1030 | 49309 |
| 1986 | 725519 | 310118 | 0.2906 | 65361 |
| 1987 | 349083 | 414524 | 0.3768 | 75295 |
| 1988 | 507432 | 473754 | 0.4816 | 92711 |
| 1989 | 496756 | 426726 | 0.3595 | 100868 |
| 1990 | 984376 | 387355 | 0.3681 | 104854 |
| 1991 | 1235351 | 345822 | 0.2466 | 109235 |
| 1992 | 773890 | 404424 | 0.2191 | 108275 |
| 1993 | 788730 | 523648 | 0.1676 | 102513 |
| 1994 | 345142 | 533841 | 0.2474 | 133753 |
| 1995 | 278752 | 487415 | 0.2526 | 125673 |
| 1996 | 972202 | 388208 | 0.2167 | 95722 |
| 1997 | 501559 | 359335 | 0.1835 | 64261 |
| 1998 | 870369 | 393129 | 0.2418 | 86849 |
| 1999 | 542287 | 382407 | 0.2214 | 92735 |
| 2000 | 612731 | 445461 | 0.2044 | 100406 |
| 2001 | 1966939 | 418931 | 0.2302 | 95352 |
| 2002 | 2533473 | 427220 | 0.2520 | 93673 |
| 2003 | 650000 | 590439 | 0.1639 | 125160 |

Table 7.5.1.1 Within survey consistency, correlation of $N$ at age a in year $y$ with $N$ at age $a+1$ in year $y+1$ over the years 1981-2004.
The numbers in bold indicate where the correlation is considered strong.

| a | $\mathrm{a}+1$ | r |
| :--- | :--- | :--- |
| 2 | 3 | $\mathbf{. 6 7}$ |
| 3 | 4 | $\mathbf{. 9 4}$ |
| 4 | 5 | $\mathbf{. 9 5}$ |
| 5 | 6 | $\mathbf{. 9 1}$ |
| 6 | 7 | $\mathbf{. 9 3}$ |
| 7 | 8 | $\mathbf{. 9 4}$ |
| 8 | 9 | $\mathbf{. 8 7}$ |
| 9 | 10 | . .77 |
| 10 | 11 | .79 |
| 11 | 12 | .64 |
| 12 | 13 | .57 |
| 13 | 14 | .68 |
| 14 | 15 | .63 |

Table 7.6.1 Icelandic summer spawners. Summary table from final amci run. Run id 20040503224235.175

## SUMMARY TABLE

| Year | Recruits <br> age 2 | SSB | $F$ <br> 5 | Catch |
| ---: | :---: | ---: | ---: | ---: |
| 1981 | 635 | 166 | 0.2705 | SOP |
| 1982 | 227 | 173 | 0.4113 | 39 |
| 1983 | 236 | 184 | 0.3604 | 56 |
| 1984 | 440 | 192 | 0.2584 | 50 |
| 1985 | 1009 | 219 | 0.2340 | 49 |
| 1986 | 587 | 234 | 0.3200 | 65 |
| 1987 | 358 | 318 | 0.3260 | 75 |
| 1988 | 518 | 358 | 0.3649 | 92 |
| 1989 | 364 | 320 | 0.3429 | 100 |
| 1990 | 878 | 294 | 0.3844 | 104 |
| 1991 | 1076 | 255 | 0.4221 | 109 |
| 1992 | 671 | 315 | 0.3946 | 108 |
| 1993 | 606 | 431 | 0.2854 | 102 |
| 1994 | 291 | 435 | 0.3958 | 133 |
| 1995 | 264 | 379 | 0.4129 | 125 |
| 1996 | 781 | 287 | 0.3601 | 95 |
| 1997 | 431 | 269 | 0.2719 | 64 |
| 1998 | 716 | 300 | 0.3510 | 86 |
| 1999 | 385 | 291 | 0.3273 | 92 |
| 2000 | 468 | 339 | 0.3255 | 100 |
| 2001 | 1556 | 305 | 0.3451 | 95 |
| 2002 | 3047 | 307 | 0.3127 | 93 |
| 2003 | 520 | 488 | 0.2496 | 125 |

Table 7.6.2 Icelandic summer spawners. Stock size.


Table 7.6.3. Icelandic summer spawning herring. Fishing mortality at age. Run id $20040503 \quad 224235.175$
Local partial fishing mortalities at age Area 1 , Fleet 1

|  | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.0006 | 0.0012 | 0.0022 | 0.0014 | 0.0011 | 0.0013 | 0.0011 | 0.0015 |
| 3 | 0.0184 | 0.0301 | 0.0459 | 0.0409 | 0.0373 | 0.0446 | 0.0383 | 0.0392 |
| 4 | 0.1067 | 0.1651 | 0.1721 | 0.1219 | 0.1169 | 0.1616 | 0.1572 | 0.1766 |
| 5 | 0.2705 | 0.4113 | 0.3604 | 0.2584 | 0.2340 | 0.3200 | 0.3261 | 0.3649 |
| 6 | 0.2705 | 0.4113 | 0.3604 | 0.2584 | 0.2340 | 0.3200 | 0.3261 | 0.3649 |
| 7 | 0.2705 | 0.4113 | 0.3604 | 0.2584 | 0.2340 | 0.3200 | 0.3261 | 0.3649 |
| 8 | 0.2705 | 0.4113 | 0.3604 | 0.2584 | 0.2340 | 0.3200 | 0.3261 | 0.3649 |
| 9 | 0.2705 | 0.4113 | 0.3604 | 0.2584 | 0.2340 | 0.3200 | 0.3261 | 0.3649 |
| 10 | 0.2705 | 0.4113 | 0.3604 | 0.2584 | 0.2340 | 0.3200 | 0.3261 | 0.3649 |
| 11 | 0.2705 | 0.4113 | 0.3604 | 0.2584 | 0.2340 | 0.3200 | 0.3261 | 0.3649 |
| 12 | 0.2705 | 0.4113 | 0.3604 | 0.2584 | 0.2340 | 0.3200 | 0.3261 | 0.3649 |
| Fref | 0.2705 | 0.4113 | 0.3604 | 0.2584 | 0.2340 | 0.3200 | 0.3261 | 0.3649 |

Local partial fishing mortalities at age
Area 1 , Fleet 1

|  | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0.0034 | 0.0060 | 0.0124 | 0.0134 | 0.0083 | 0.0140 | 0.0181 | 0.0128 |
| 3 | 0.0404 | 0.0477 | 0.0679 | 0.0708 | 0.0548 | 0.1084 | 0.1155 | 0.0971 |
| 4 | 0.1635 | 0.1903 | 0.2118 | 0.2081 | 0.1637 | 0.2326 | 0.2710 | 0.2366 |
| 5 | 0.3429 | 0.3844 | 0.4221 | 0.3946 | 0.2854 | 0.3958 | 0.4129 | 0.3601 |
| 6 | 0.3429 | 0.3844 | 0.4221 | 0.3946 | 0.2854 | 0.3958 | 0.4129 | 0.3601 |
| 7 | 0.3429 | 0.3844 | 0.4221 | 0.3946 | 0.2854 | 0.3958 | 0.4129 | 0.3601 |
| 8 | 0.3429 | 0.3844 | 0.4221 | 0.3946 | 0.2854 | 0.3958 | 0.4129 | 0.3601 |
| 9 | 0.3429 | 0.3844 | 0.4221 | 0.3946 | 0.2854 | 0.3958 | 0.4129 | 0.3601 |
| 10 | 0.3429 | 0.3844 | 0.4221 | 0.3946 | 0.2854 | 0.3958 | 0.4129 | 0.3601 |
| 11 | 0.3429 | 0.3844 | 0.4221 | 0.3946 | 0.2854 | 0.3958 | 0.4129 | 0.3601 |
| 12 | 0.3429 | 0.3844 | 0.4221 | 0.3946 | 0.2854 | 0.3958 | 0.4129 | 0.3601 |
|  |  |  |  |  |  |  |  |  |
| Fref | 0.3429 | 0.3844 | 0.4221 | 0.3946 | 0.2854 | 0.3958 | 0.4129 | 0.3601 |

Local partial fishing mortalities at age Area 1 , Fleet 1
$\left.\begin{array}{rrrrrrrr} & & 1997 & 1998 & 1999 & 2000 & 2001 & 2002\end{array}\right)$

Table 7.7.1 Input data for short term prediction.
MFDP version 1
Run: a
Time and date: 00:40 5/4/2004
Fbar age range: 5-12


Table 7.7.2 Short term prediction table.
MFDP version 1
Run: b
MFDP Index file 3/05/2004
Time and date: 00:16 5/4/2004
Fbar age range: 5-12

| $2004 / 2005$ |  |  |  |  |
| :---: | :---: | :---: | :---: | ---: |
| Biomass | SSB | FMult | FBar | Landings |
| 684 | 525 | 0.8814 | 0.22 | 106 |


| 2005/2006 |  |  |  | 2006/2007 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 668 | 530 | 0 | 0 | 0 | 753 | 614 |
| - | 530 | 0.1 | 0.025 | 13 | 739 | 601 |
| . | 530 | 0.2 | 0.0499 | 26 | 725 | 588 |
| . | 530 | 0.3 | 0.0749 | 39 | 711 | 575 |
| . | 530 | 0.4 | 0.0998 | 51 | 698 | 563 |
| . | 530 | 0.5 | 0.1248 | 64 | 685 | 551 |
| . | 530 | 0.6 | 0.1498 | 75 | 673 | 539 |
| - | 530 | 0.7 | 0.1747 | 87 | 660 | 527 |
| . | 530 | 0.8 | 0.1997 | 98 | 648 | 516 |
| . | 530 | 0.9 | 0.2246 | 109 | 637 | 505 |
| . | 530 | 1 | 0.2496 | 120 | 625 | 494 |
| . | 530 | 1.1 | 0.2746 | 131 | 614 | 484 |
| . | 530 | 1.2 | 0.2995 | 141 | 603 | 473 |
| . | 530 | 1.3 | 0.3245 | 151 | 592 | 463 |
| . | 530 | 1.4 | 0.3494 | 161 | 582 | 454 |
| . | 530 | 1.5 | 0.3744 | 171 | 571 | 444 |
| . | 530 | 1.6 | 0.3994 | 180 | 561 | 435 |
| . | 530 | 1.7 | 0.4243 | 189 | 552 | 426 |
| . | 530 | 1.8 | 0.4493 | 198 | 542 | 417 |
| - | 530 | 1.9 | 0.4742 | 207 | 533 | 408 |
| . | 530 | 2 | 0.4992 | 215 | 524 | 400 |

Input units are millions and kg - output in kilotonnes


Figure 7.2.1 Icelandic summer spawning herring. Catch at age in numbers in 2003.


Figure 7.3.1 Icelandic summer spawning herring. Acoustic estimates by age in millions.


Figure 7.4.1 Results from the assessment made in 2003 and the update assessment (2004).


Figure 7.5.1.1. Catchcurves by yearclasses and years. Grey lines correspond to $\mathrm{Z}=0.4$


Figure 7.5.1.1 cont. Catchcurves by yearclasses and years. Grey lines correspond to $\mathrm{Z}=0.4$


Figure 7.5.1.2. Catchcurves by yearclasses and years. Grey lines correspond to $\mathrm{Z}=0.4$


Figure 7.5.1.2 cont.. Catchcurves by yearclasses and years. Grey lines correspond to $\mathrm{Z}=0.4$


Figure 7.5.1.3. Catchcurves made from acoustic survey data. Grey lines correspond to $\mathrm{Z}=0.4$

age

Figure 7.5.1.3 cont. Catchcurves made from acoustic survey data. Grey lines correspond to $\mathrm{Z}=0.4$


Figure 7.5.1.4. Catchcurves from survey data. Grey lines correspond to $\mathrm{Z}=0.4$.

year
Figure 7.5.1.4 cont. Catchcurves from survey data. Grey lines correspond to $\mathrm{Z}=0.4$.


Figure 7.5.1.5. Comparison between ages in the survey. Yearclasses are denoted in the graphs. Solid line is the fitted line through origin and dotted line is line with slope 1.


Figure 7.5.1.6 A biomass index for the icelandic acoustic survey 1981-2004 by age groups 3-9 and 3-11

Run 1: Tuning year range 1981-2004


Run 2: Tuning
year range 1988-2004


Run 3: Tuning year range 1988-2004 - only age 3 and 4


Figure 7.5.2.1 Retrospective patterns in $q$ in Adapt runs.

Run 1


Run 2


Run 3


Figure 7.5.2.2 Retrospective analysis of SSB

## Run 1



Run 2


## Run 3



Figure 7.5.2.3 Retrospective analysis of fishing mortality (F5-10 unweighted)


Figure 7.5.4.1 Diagnostic plots for run 1.


Figure 7.5.4.2 Retrospective run for amci run 1.


Figure 7.5.5.1 Diagnostic plots from a run with Adapt in AD-Modelbuilder.


Figure 7.5.5.2 A retrospective plot for an Adapt run.


Figure 7.5.6.1 Comparisons of exploratory runs and the assessment made last year (2003).

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## ANNEX I

## Northern Pelagic and Blue Whiting Fisheries Working Group

ICES, Headquarters, 27 April - 4 May 2004

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## ANNEX II

```
DETAILS OF DATA FILLING-IN
```



Filling-in for record : (127) Germany 3 IIa Mean Weighted by Sampled Catches of:
$\gg(3)$ Iceland 3 IIa
$\gg$ ( 23) Faroe Islands 3 IIa
$\gg$ (155) Norway 3 IIa
$\gg$ (183) Russia 3 IIa

Filling-in for record : (205) Denmark 1 IIa Mean Weighted by Sampled Catches of:
>> (153) Norway 1 IIa
>> (181) Russia 1 IIa
Filling-in for record : (206) Denmark 2 IIa Mean Weighted by Sampled Catches of:
$\gg(2)$ Iceland 2 IIa
$\gg$ ( 22) Faroe Islands 2 IIa
$\gg$ (154) Norway 2 IIa
>> (182) Russia 2 IIa
Filling-in for record : (208) Denmark 4 IIa
Using Only
>> (184) Russia 4 IIa
Filling-in for record : (258) Sweden 2 IIa
Mean Weighted by Sampled Catches of:
>> ( 2) Iceland 2 IIa
$\gg(22)$ Faroe Islands 2 IIa
$\gg$ (154) Norway 2 IIa
>> (182) Russia 2 IIa
Filling-in for record : (259) Sweden 3 IIa
Mean Weighted by Sampled Catches of:
>> ( 3) Iceland 3 IIa
>> ( 23) Faroe Islands 3 IIa
>> (155) Norway 3 IIa
>> (183) Russia 3 IIa
Filling-in for record : (185) Russia 1 IIb
Using Only
>> (177) Russia 1 I
Filling-in for record : (186) Russia
Using Only
>> (177) Russia 1 I


```
    Filling-in for record : (222) Denmark
    2 Vb
Mean Weighted by Sampled Catches of:
    >> ( 10) Iceland 
    Filling-in for record : (223) Denmark
3 Vb
Mean Weighted by Sampled Catches of:
    >> ( 11) Iceland 3 Vb
    >> ( 31) Faroe Islands 3 Vb
    Filling-in for record : (224) Denmark
    Vb
Mean Weighted by Sampled Catches of:
    >> ( 12) Iceland 4 Vb
    >> ( 32) Faroe Islands 4 Vb
    Filling-in for record : (239) France
Mean Weighted by Sampled Catches of:
    >> ( 11) Iceland 3 Vb
    >> ( 31) Faroe Islands 3 Vb
    Filling-in for record : (240) France
4 Vb
Mean Weighted by Sampled Catches of
    >> ( 12) Iceland 4 Vb
    >> ( 32) Faroe Islands 4 Vb
    Filling-in for record : ( 69) The Netherlands
1 VIa
Mean Weighted by Sampled Catches of:
    >> ( 37) Faroe Islands 1 VIb
    >> (165) Norway 1 VIb
    >> (193) Russia 1 VIb
    Filling-in for record : (105) Scotland 1 VIa
Mean Weighted by Sampled Catches of:
    >> ( 37) Faroe Islands 1 VIb
    >> (165) Norway 1 VIb
    >> (193) Russia 1 VIb
    Filling-in for record : (106) Scotland 2 VIa
Mean Weighted by Sampled Catches of:
    >> ( 70) The Netherlands 2 VIa
    >> (162) Norway 2 VIa
    Filling-in for record : (108) Scotland 4 VIa
Using Only
    >> (163) Norway 3 VIa
    Filling-in for record : (227) Denmark 3 VIa
Using Only
    >> (163) Norway 3 VIa
    Filling-in for record : (228) Denmark 4 VIa
Using Only
    >> (163) Norway 3 VIa
    Filling-in for record : ( 13) Iceland 1 VIb
Mean Weighted by Sampled Catches of:
    >> ( 37) Faroe Islands 1 VIb
    >> (165) Norway 1 VIb
    >> (193) Russia 1 VIb
    Filling-in for record : ( 14) Iceland 2 VIb
Mean Weighted by Sampled Catches of:
    >> ( 38) Faroe Islands 2 VIb
```

```
>> (166) Norway 2 VIb
>> (194) Russia 2 VIb
```

Filling-in for record : ( 73) The Netherlands 1 VIb
Mean Weighted by Sampled Catches of:

| $\gg$ | $(37)$ | Faroe Islands | 1 VIb |
| :--- | :--- | :--- | :--- |
| $\gg$ | $(165)$ | Norway | 1 VIb |
| $\gg$ | $(193)$ | Russia | 1 VIb |

Filling-in for record : (109) Scotland
1 VIb
Mean Weighted by Sampled Catches of:
$\gg$ ( 37) Faroe Islands 1 VIb
$\gg$ (165) Norway 1 VIb
$\gg$ (193) Russia 1 VIb
Filling-in for record : (110) Scotland 2 VIb
Mean Weighted by Sampled Catches of :
$\gg$ (38) Faroe Islands 2 VIb
$\gg$ (166) Norway 2 VIb
$\gg$ (194) Russia 2 VIb
Filling-in for record : (141) Germany 1 VIb
Mean Weighted by Sampled Catches of:

| $\gg$ | $(37)$ | Faroe Islands | 1 VIb |
| :--- | :--- | :--- | :--- |
| $\gg$ | $(165)$ | Norway | 1 VIb |
| $\gg$ | $(193)$ | Russia | 1 VIb |

    Filling-in for record : (142) Germany
    2 VIb
Mean Weighted by Sampled Catches of:
$\gg$ ( 38) Faroe Islands 2 VIb
$\gg(166)$ Norway 2 VIb
$\gg$ (194) Russia 2 VIb
Filling-in for record : (195) Russia 3 VIb
Mean Weighted by Sampled Catches of:
$\gg$ (38) Faroe Islands 2 VIb
$\gg(166)$ Norway 2 VIb
$\gg$ (194) Russia 2 VIb
Filling-in for record : (241) France 1 VIb
Mean Weighted by Sampled Catches of:
$\gg$ ( 37) Faroe Islands 1 VIb
$\gg$ (165) Norway 1 VIb
$\gg$ (193) Russia 1 VIb
Filling-in for record : (113) Scotland
1 VIIb
Mean Weighted by Sampled Catches of:
>> ( 41) Faroe Islands 1 VIIc
$\gg$ (77) The Netherlands 1 VIIc
$>$ (173) Norway 1 VIIbc
Filling-in for record : (145) Germany 1 VIIbc
Mean Weighted by Sampled Catches of:
$\gg$ ( 41) Faroe Islands 1 VIIc
$\gg$ (77) The Netherlands 1 VIIc
$\gg$ (173) Norway 1 VIIbc
Filling-in for record : (245) France 1 VIIbc
Mean Weighted by Sampled Catches of:

| $\gg$ | $(41)$ | Faroe Islands | 1 VIIc |
| :--- | :--- | :--- | :--- |
| $\gg$ | $(77)$ | The Netherlands | 1 VIIc |
| $\gg$ | $(173)$ | Norway | 1 VIIbc |

    Filling-in for record : ( 78) The Netherlands 2 VIIc
    Mean Weighted by Sampled Catches of:

```
>> ( 41) Faroe Islands 1 VIIc
>> ( 77) The Netherlands 1 VIIc
>> (173) Norway 1 VIIbc
```

```
Filling-in for record : (117) Scotland
1 VIIc
```

Mean Weighted by Sampled Catches of:
>> ( 41) Faroe Islands 1 VIIc
>> ( 77) The Netherlands 1 VIIc
>> (173) Norway 1 VIIbc
Filling-in for record : ( 17) Iceland

1 VIIgk+XII

1 XII
Mean Weighted by Sampled Catches of:

```
>> ( 49) Faroe Islands 1 VIIgk+XII
    >> (169) Norway 1 VIIgk
    >> (197) Russia 1 VIIgk+XII
```

    Filling-in for record : (198) Russia 2 VIIgk+XII
    Mean Weighted by Sampled Catches of:
$\gg$ ( 41) Faroe Islands 1 VIIc
>> ( 77) The Netherlands 1 VIIc
>> (173) Norway 1 VIIbc
Filling-in for record : (249) France 1 VIIg-k
Mean Weighted by Sampled Catches of:
>> ( 49) Faroe Islands 1 VIIgk+XII
>> (169) Norway 1 VIIgk
>> (197) Russia 1 VIIgk+XII
Filling-in for record : ( 81) The Netherlands 1 VIIh
Mean Weighted by Sampled Catches of:
>> ( 49) Faroe Islands 1 VIIgk+XII
>> (169) Norway 1 VIIgk
>> (197) Russia 1 VIIgk+XII
Filling-in for record : ( 84) The Netherlands
Using Only
>> ( 95) The Netherlands 3 VIIIabd
Filling-in for record : ( 93) The Netherlands 1 VIIIabd
Using Only
>> ( 95) The Netherlands 3 VIIIabd
Filling-in for record : ( 94) The Netherlands 2 VIIIabd
Using Only
>> ( 95) The Netherlands 3 VIIIabd
Filling-in for record : (255) France 3 VIIIabd
Using Only
>> ( 95) The Netherlands 3 VIIIabd
Filling-in for record : ( 85) The Netherlands 1 VIIj
Mean Weighted by Sampled Catches of:

```
>> ( 49) Faroe Islands 1 VIIgk+XII
    >> (169) Norway 1 VIIgk
    >> (197) Russia 1 VIIgk+XII
```

```
    Filling-in for record : ( 87) The Netherlands
Using Only
    >> ( 86) The Netherlands 2 VIIj
    Filling-in for record : (121) Scotland 1 VIIj
Mean Weighted by Sampled Catches of:
    >> ( 49) Faroe Islands 1 VIIgk+XII
    >> (169) Norway 1 VIIgk
    >> (197) Russia 1 VIIgk+XII
    Filling-in for record : (149) Germany 1 VIIj
Mean Weighted by Sampled Catches of:
    >> ( 49) Faroe Islands 1 VIIgk+XII
    >> (169) Norway 1 VIIgk
    >> (197) Russia 1 VIIgk+XII
    Filling-in for record : ( 89) The Netherlands
1 VIIk
Mean Weighted by Sampled Catches of:
    >> ( 49) Faroe Islands 1 VIIgk+XII
    >> (169) Norway 1 VIIgk
    >> (197) Russia 1 VIIgk+XII
    Filling-in for record : (277) Ireland 1 VIa
Mean Weighted by Sampled Catches of:
    >> ( 37) Faroe Islands 1 VIb
    >> (165) Norway 1 VIb
    >> (193) Russia 1 VIb
    Filling-in for record : (281) Ireland 1 VIb
Mean Weighted by Sampled Catches of:
    >> ( 37) Faroe Islands 1 VIb
    >> (165) Norway 1 VIb
    >> (193) Russia 1 VIb
    Filling-in for record : (285) Ireland
                                    1 VIIb
Mean Weighted by Sampled Catches of:
    >> ( 41) Faroe Islands 1 VIIc
    >> ( 77) The Netherlands 1 VIIc
    >> (173) Norway 1 VIIbc
    Filling-in for record : (286) Ireland 2 VIIb
Mean Weighted by Sampled Catches of:
    >> ( 41) Faroe Islands 1 VIIc
    >> ( 77) The Netherlands 1 VIIc
    >> (173) Norway 1 VIIbc
    Filling-in for record : (305) Ireland 1 XII
Mean Weighted by Sampled Catches of:
    >> ( 49) Faroe Islands 1 VIIgk+XII
    >> (169) Norway 1 VIIgk
    >> (197) Russia 1 VIIgk+XII
    Filling-in for record : (294) Ireland 2 VIIg
Mean Weighted by Sampled Catches of:
    >> ( 41) Faroe Islands 1 VIIc
    >> ( 77) The Netherlands 1 VIIc
    >> (173) Norway 1 VIIbc
    Filling-in for record : (297) Ireland 1 VIIj
Mean Weighted by Sampled Catches of:
\begin{tabular}{lll}
\(\gg\) & \((49)\) & Faroe Islands \\
\(\gg(169)\) & Norway & 1 VIIgk+XII \\
\(\gg(197)\) & Russia & 1 VIIgk \\
\(\gg\) & 1 VIIgk+XII
\end{tabular}
```

```
    Filling-in for record : (298) Ireland 2 VIIj
Using Only
    >> ( 86) The Netherlands 2 VIIj
    Filling-in for record : (300) Ireland 4 VIIj
Using Only
    >> (301) Ireland 1 VIIk
    Filling-in for record : (261) Sweden 1 IIIa
Using Only
    >> (210) Denmark 2 IIIa
    Filling-in for record : (262)
Using Only
    >> (211) Denmark
    Filling-in for record : (264)
Using Only
    >> (211) Denmark
    Filling-in for record : (267)
Using Only
    >> (215) Denmark
    Filling-in for record : (268)
Using Only
    >> (215) Denmark 3 IVa
    Filling-in for record : (309)
Using Only
    >> (317) Norway-mix
    Filling-in for record : (310)
Using Only
    >> (323) Norway-mix 3 IVa
    Filling-in for record : (311)
Using Only
    >> (324) Norway-mix 4 IVa
    Filling-in for record : (312) Faroe Islands-mix 1 IVa
Using Only
    >> (321) Norway-mix 1 IVa
    Filling-in for record : (313) Faroe Islands-mix 2 IVa
Using Only
    >> (322) Norway-mix 2 IVa
    Filling-in for record : (314) Faroe Islands-mix 3 IVa
Using Only
    >> (323) Norway-mix 3 IVa
    Filling-in for record : (315)
Using Only
    >> (324) Norway-mix
    Filling-in for record : (316)
Using Only
    >> (323) Norway-mix
    Filling-in for record : (320)
Using Only
    >> (319) Norway-mix 3 IIa
```


## APPENDIX

## Special opinion of Russian delegation

In order to give managers more information about different perception of stock size and fishery perspectives, Russian delegation conceders it to be important to present the results of short-term forecast, based on the results of stock assessment by means of ISVPA and agreed by the WG parameters on weight-at-age, natural mortality and maturity-atage in prognostic years.

Blue whiting:

| 2004 |  |  |  |  | 2005 |  |  |  |  | 2006 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 8493 | 5020 | 1 | 0.53 | 2365 | 6710 | 4680 | 0.0 | 0.000 | 0 | 7367 | 5305 |
|  |  |  |  |  |  | 4626 | 0.1 | 0.053 | 241 | 7113 | 5089 |
|  |  |  |  |  |  | 4574 | 0.2 | 0.106 | 471 | 6870 | 4883 |
|  |  |  |  |  |  | 4522 | 0.3 | 0.159 | 691 | 6638 | 4687 |
|  |  |  |  |  |  | 4471 | 0.4 | 0.212 | 902 | 6417 | 4500 |
|  |  |  |  |  |  | 4420 | 0.5 | 0.265 | 1103 | 6205 | 4322 |
|  |  |  |  |  |  | 4370 | 0.6 | 0.318 | 1295 | 6003 | 4152 |
|  |  |  |  |  |  | 4321 | 0.7 | 0.371 | 1479 | 5810 | 3990 |
|  |  |  |  |  |  | 4272 | 0.8 | 0.424 | 1656 | 5625 | 3835 |
|  |  |  |  |  |  | 4224 | 0.9 | 0.477 | 1824 | 5448 | 3687 |
|  |  |  |  |  |  | 4176 | 1.0 | 0.530 | 1986 | 5279 | 3546 |

In forecas
Fbar: 3-7

Norwegian spring-spawning herring:

| 2004 |  |  |  |  | 2005 |  |  |  |  | 2006 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Biomass | SSB | FMult | FBar | Landings | Biomass | SSB | FMult | FBar | Landings | Biomass | SSB |
| 13374 | 10768 | 1.54 | 0.101 | 825 | 12612 | 11089 | 0.0 | 0.000 | 0 | 12211 | 11274 |
|  |  |  |  |  | . | 11074 | 0.1 | 0.007 | 66 | 12139 | 11204 |
|  |  |  |  |  | . | 11059 | 0.2 | 0.014 | 132 | 12067 | 11134 |
|  |  |  |  |  | . | 11044 | 0.3 | 0.022 | 197 | 11996 | 11065 |
|  |  |  |  |  | . | 11030 | 0.4 | 0.029 | 261 | 11925 | 10997 |
|  |  |  |  |  | . | 11015 | 0.5 | 0.036 | 325 | 11855 | 10929 |
|  |  |  |  |  | . | 11000 | 0.6 | 0.043 | 389 | 11785 | 10862 |
|  |  |  |  |  | . | 10986 | 0.7 | 0.050 | 452 | 11716 | 10795 |
|  |  |  |  |  |  | 10971 | 0.8 | 0.057 | 514 | 11648 | 10729 |
|  |  |  |  |  | . | 10956 | 0.9 | 0.065 | 576 | 11580 | 10664 |
|  |  |  |  |  | . | 10942 | 1.0 | 0.072 | 638 | 11513 | 10599 |
|  |  |  |  |  |  | 10927 | 1.1 | 0.079 | 699 | 11446 | 10534 |
|  |  |  |  |  | . | 10913 | 1.2 | 0.086 | 759 | 11380 | 10470 |
|  |  |  |  |  | . | 10898 | 1.3 | 0.093 | 819 | 11314 | 10406 |
|  |  |  |  |  |  | 10884 | 1.4 | 0.100 | 879 | 11249 | 10343 |
|  |  |  |  |  | . | 10869 | 1.5 | 0.108 | 938 | 11184 | 10281 |
|  |  |  |  |  | . | 10855 | 1.6 | 0.115 | 996 | 11120 | 10219 |
|  |  |  |  |  |  | 10840 | 1.7 | 0.122 | 1054 | 11056 | 10157 |
|  |  |  |  |  |  | 10834 | 1.740 | 0.125 | 1077 | 11031 | 10133 |
|  |  |  |  |  | . | 10826 | 1.8 | 0.129 | 1112 | 10993 | 10096 |
|  |  |  |  |  | - | 10811 | 1.9 | 0.136 | 1169 | 10931 | 10036 |
|  |  |  |  |  | . | 10797 | 2.0 | 0.144 | 1226 | 10868 | 9976 |

Recruitment: GM 1950-2002 = 9794 millions

It seems expedient also to attract attention to the fact that for blue whiting stock assessment the model which gave marginally low stock estimates among the 4 models tested at the Working Group was chosen.

Russian delegation underlines also that shortened duration of the Working Group meeting restricts possibilities for comprehensive scientific discussions and propose to restore duration of the Working Group to 10 days.

# Review of Northern Pelagic and Blue Whiting WG report 2004. <br> 26-27 May, 2004 ICES, Copenhagen <br> Technical minutes 

## Participants:

Jan Horbowy, sub-group chair, Poland
Andrē Forest, reviewer, France
John Simmonds, reviewer, UK
Asta Gusmundsdotir, WG chair, Island

## Terms of reference for NPBWWG

a) assess the status of and provide catch options for 2005 for the Norwegian spring-spawning herring stock
b) assess the status of and provide catch options for 2005 for the blue whiting stock;
c) assess the status of and provide catch options for the 2004-2005 season for the Icelandic summerspawning herring stocks
d) assess the status of capelin in Subareas V and XIV and provide catch options for the summer/autumn 2004 and winter 2005 seasons
e) provide as detailed information as possible on the age/size composition in different segments of the blue whiting fishery
f) provide information on the species compositions in those fisheries that take appreciable amounts of blue whiting, and on the age/size composition by species of these catches [EC request for information on the industrial fisheries]
g) propose measures to reduce exploitation of blue whiting juveniles and evaluate the potential effect on the stock and the fisheries. The evaluation should include, but not be restricted to the effects of introducing a minimum size and closed areas/seasons;
h) continue the evaluation of candidates of harvest control rules for blue whiting
i) provide specific information on possible deficiencies in the 2004 assessments including, at least, any major inadequacies in the data on catches, effort or discards; any major inadequacies in research vessel surveys data, and any major difficulties in model formulation, including inadequacies in available software. The consequences of these deficiencies for the assessment of the status of the stocks and for the projection should be clarified
j) comment on this meeting's assessments compared to the last assessment of the same stock, for stocks for which a full or update assessment is presented
k) document fully the methods to be applied in subsequent update assessments and list factors that would warrant reconsideration of doing an update, and consider doing a benchmark ahead of schedule, for stocks for which benchmark assessments are done.

The WG has completed TOR a-e and i dealing with all the assessment of stocks, TOR f on blue whiting fisheries is completed. TOR $g$ was carried and reported in last years WG report ICES CM2003/ACFM23.

Three TOR were not addressed,
f, was not completed due lack of time and incomplete age data,
h was not attempted due to shortage of appropriate software;
k. The two major stocks are currently under review of a study group to determine benchmark assessments once these are complete this will form the basis for further benchmark assessments.

## Ecological Effects.

The WG provided a good overview of the influence of the environment on the development and distribution of stocks in this area. They provided model based estimates of recruitment. While most of the information used came from survey indices of abundance of recruits some additional information was obtained from environmental data. The WG should consider these models at the start of the meeting next year and decide if the proposed methods should be used in future estimates of recruitment.

## Comments by Stock

## Norwegian Spring Spawning herring

## Data conformity

According to a request from last year review group information pertinent to the assessment of this stock are available on the ICES server and not only at a website in Norway.

Sop of Icelandic catch in area IIa is high at 1.49. Catch total was confirmed as correct. Could not be resolved at the group, the influence was determined to be negligible.
Maturity at age used in the assessment were seen to be inconsistent, in particular the ages 34 and 5 in year 1969 through 1973.

The WG indicated there were density dependent growth and maturity effects to be taken into account in the projections. However, year independent values are used in the assessment since 1993. This inconsistent use of data should be resolved

Mean weight at age in the period 1972 to 1992 show increase with age until 10 or 11 year then decreases are found. Could this be validated.

## Assessments

As proposed by the WG the reviewers endorse the view that of the available assessments the Sea Star assessment is the best assessment for this stock.

Sea Star.
Error distribution assumptions (differing among data sets) are included in the model, these are not supported by tables or graphs in this year's report, these should be included each year so that it can be shown that the error distribution choice is appropriate.
It is suggested to include in the model also negative binomial error structure, which appears to work well with havy tailed survey data.

A standardised simple residual plot between observations and model estimates for all data should be provided. Some graphs for some surveys were included but a coherent residual plot is required.
Extensive range of different types of data are included in the model, It seems unlikely that all sources are equally informative, there is no information on weighting provided, Could this be provided in future and if possible different weighting of different data in the assessment explored. In particular the recruit indices may be informative in differing amounts, the wide span of ages from both catch and surveys have differing errors with age and it may be helpful to account for this with weighting.

There is a general tendency in figures produced for Sea Star to omit labels and units on the axes of graphs. Please ensure they are fully labelled to avoid confusion.

## ISVPA

Its not possible to determine how survey catchability is dealt with within the ISVPA as used for this stock. In future could the age dependent or independent catchabilities be included.

Figure 3.5.3 shows a 'de trended' ISVPA fit. This is not explained in the report.
There are changes in the choice of model settings in comparison with those used in the previous assessment (minimisation of MDN except SSE for survey N5 in the 2003 assessment, compared to minimisation of MDN except AMD for surveys N3 and N4 in the 2004 assessment). The effect of these
are not explained and the influence of these changes are not shown. This makes it difficult to judge the effect and importance of these changes and therefore the utility of the model.

General
ISVPA produces a rising selection function for catch at age from 1 to the older ages in the recent period. This is in contrast to the pattern from Sea Star which rises to 7 or 8 years and is domed or flatter for older herring. This difference is thought to be important for the assessment but is not discussed. The effect of the elevated part of the selection pattern (8-11 years) should be examined in future.

The surveys are documented intermittently by survey 'Name', or number, please ensure that the number is used on all descriptions tables and figures, so there is one consistent key to identify each survey.

The assessment is strongly dependent on the use of the 'feeding area survey'. A different treatment of this survey in ISVPA was the main cause of the difference in the 'ISVPA' and 'Sea Star' assessments. Although the fit was shown to be better in ISVPA with this change there was no discussion of why only this survey should be treated this way and why this might be the preferred method. Thus it was not possible to make a clear choice of which was the correct use. Under these circumstances the reviewers concluded that the method used last year for both ISVPA and Sea Star should be followed in preference to the new unexplained use taken this year by in ISVPA.

## Projections

Short projections included maturity and weight at age that varied with year. It is good to see that this is being considered for projections for this stock but the values used were not fully documented in the table provided.

The mean weights at age in the projections seem inconsistent for ages, particularly for age 3 . Its not expected that this will influence the SSB or the estimated catch used for management by much, however, the validity of these weights should be checked.

The short term predictions were rerun during the review and very small changes were found, these were include in the ACFM summary sheets. The new file is in the review group directory under 'Asta'.

In 2004 a TAC constraint was used for the projections. While the reviews support the use of this method, no explanation to support this choice was documented.

## Capelin in the Icelandic-East Greenland - Jan Mayen Area

## Data conformity

Its not clear from the report if the Capelin Acoustic Surveys are used as absolute or relative indices of abundance. If they are used as absolute in management, the Target Strength of Capelin used and its origin should be quoted. Perhaps validation of the survey against subsequent catch could be used to improve certainty with the method.

## Assessments/ Projections

The assessment / projections rely on a constant time invariant natural mortality. Particularly of surveys from different times of year are to be used in management (as seems to the case recently) the validity of this assumption this needs to be considered.

## Blue Whiting

## Data conformity

This year the Norwegian survey of the spawning grounds alone was used in the assessment. The more extensive 4 ship survey was available. Unless there are reasons to believe that the additional data is incorrect it would be preferable to use the higher effort survey, maybe initially only from a standardised area for consistency. It is appreciated that it takes considerable work and organisation to get the survey data from such a survey combined in time for the WG.

## Assessments

As proposed by the WG the reviewers endorse the view that of the available assessments the AMCI assessment is the best assessment for this stock.

The WG should be commended for trying 4 different packages for assessments for this stock. The broad correspondence between the results provides good support for the final assessment.

Comparing the assessments coming from the 4 models gave rise to some concerning patterns of residuals, particularly in ISVPA and SMS. Positive and Negative cohort dependent residuals were seen in both these models but in different directions in the different models for some of the older ages currently in the stock. If the causes of these effects could be established it may be possible to determine more easily which model is providing the better representation of the stock.

## Projections

The methodology used for stochastic stock projections for this stock looks reasonable and the results do not look inconsistent. However, this software has not been validated against other software. It would be useful if this could be done before the next NPBWWG. One method would be to set variability to zero and to compare with deterministic software.

## Icelandic summer spawning herring

Data conformity
The assessment relies heavily on the acoustic survey results. This survey is the result of several areas combined. When changes are occurring and there are both increases in numbers of juveniles and decreases in adults seen in these surveys it would be helpful to have a timeseries of area based values for comparison. This might allow such changes to be better supported.

## Assessments

The reviews accepted the assessment of this stock.
There are some concerns that the assessment has a strong retrospective behaviour. This effect was reduced when fewer ages were used in the fit. This effect should be evaluated further. It may be due to ageing errors, increased variance of estimates with age and flat weighting in the assessment model (or something else). Nevertheless it should be investigated further.

## Projections

Projections for 2 years ahead were provided. However, only one year ahead is needed as the stock is managed in year. Appropriate tables should be provided in the future.

The section on management advice was not fully completed. This made filling out the summary sheets more difficult

## Norwegian spring-spawning herring

## Input to the short term prediction

Landings in $2004 \quad .825$ million tonnes
Fbar age range: 5-14, Fbar is weighted with population numbers January 1
Spawning stock biomass in $2004 \quad 6.96595$ million tonnes Fbar in 2004, 0.166646
Basis for weight in stock prediction:condition factor Basis for weight in catch prediction:condition factor Part of year before spawning: . 1

| Age | Numbers <br> (billion) | Weight <br> stock <br> $(\mathrm{kg})$ | Weight <br> stock <br> $(\mathrm{kg})$ | Weight <br> catch <br> $(\mathrm{kg})$ | Fraction <br> mature | Fraction <br> mature | Fraction <br> mature | Exploita- <br> tion |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 2004 | 2004 | 2005 |  | 2004 | 2005 | 2006 | pattern |
| 0 | 0.000 | 0.001 | 0.001 | 0.007 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1 | 69.033 | 0.010 | 0.010 | 0.063 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 | 33.318 | 0.037 | 0.037 | 0.094 | 0.000 | 0.000 | 0.000 | 0.002 |
| 3 | 1.612 | 0.055 | 0.055 | 0.171 | 0.000 | 0.150 | 0.000 | 0.104 |
| 4 | 0.629 | 0.212 | 0.214 | 0.218 | 0.300 | 0.300 | 0.450 | 0.031 |
| 5 | 9.479 | 0.241 | 0.243 | 0.257 | 0.900 | 0.900 | 0.900 | 0.068 |
| 6 | 9.949 | 0.279 | 0.281 | 0.288 | 1.000 | 1.000 | 1.000 | 0.171 |
| 7 | 0.898 | 0.302 | 0.304 | 0.319 | 1.000 | 1.000 | 1.000 | 0.154 |
| 8 | 0.958 | 0.337 | 0.340 | 0.327 | 1.000 | 1.000 | 1.000 | 0.201 |
| 9 | 0.099 | 0.354 | 0.357 | 0.354 | 1.000 | 1.000 | 1.000 | 0.248 |
| 10 | 0.253 | 0.355 | 0.358 | 0.363 | 1.000 | 1.000 | 1.000 | 0.296 |
| 11 | 0.596 | 0.360 | 0.363 | 0.361 | 1.000 | 1.000 | 1.000 | 0.237 |
| 12 | 1.995 | 0.371 | 0.374 | 0.381 | 1.000 | 1.000 | 1.000 | 0.188 |
| 13 | 1.000 | 0.400 | 0.403 | 0.397 | 1.000 | 1.000 | 1.000 | 0.179 |
| 14 | 0.208 | 0.412 | 0.415 | 0.402 | 1.000 | 1.000 | 1.000 | 0.176 |
| 15 | 0.044 | 0.445 | 0.448 | 0.450 | 1.000 | 1.000 | 1.000 | 0.174 |
| 16 | 0.283 | 0.445 | 0.448 | 0.450 | 1.000 | 1.000 | 1.000 | 0.174 |


| Short term prediction |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| SSB | Fbar | Landings | Biomass | SSB |
| 2005 |  | 2005 | 2006 | 2006 |
| 6.317 | 0.084 | 0.605 | 8.702 | 6.593 |
| 6.307 | 0.101 | 0.721 | 8.585 | 6.483 |
| 6.296 | 0.118 | 0.835 | 8.470 | 6.374 |
| 6.291 | 0.125 | 0.884 | 8.421 | 6.327 |
| 6.286 | 0.134 | 0.947 | 8.357 | 6.267 |
| 6.275 | 0.151 | 1.057 | 8.246 | 6.162 |
| 6.265 | 0.168 | 1.166 | 8.137 | 6.059 |
| 6.254 | 0.185 | 1.273 | 8.029 | 5.958 |
| 6.244 | 0.202 | 1.378 | 7.923 | 5.859 |
| 6.233 | 0.218 | 1.482 | 7.818 | 5.762 |
| 6.223 | 0.235 | 1.584 | 7.716 | 5.666 |
| 6.212 | 0.252 | 1.684 | 7.614 | 5.572 |


[^0]:    ${ }^{1}$ Preliminary, as provided by Working Group members.

[^1]:    * Preliminary
    ** Predicted

[^2]:    ${ }^{1}$ AMCI 2.3 was used for scanning over terminal F's.
    ${ }^{2}$ Age 1 was down-weighted by factor 0.01 .

[^3]:    ) Including some directed fishery also in Division IVa.
    2) Revised for the years 1987, 1988, 1989, 1992, 1995,1996,1997

[^4]:    $\square$ used in the assessment

[^5]:    * Predicted

