International Council for the Exploration of the Sea
Conseil International pour
l'Exploration de la Mer

# Report on the Pandalus Assessment Working Group 

27 October - 5 November 2004
ICES Headquarters, Copenhagen, Denmark

This report is not to be quoted without prior consultation with the General Secretary. The document is a report of an Expert Group under the auspices of the International Council for the Exploration of the Sea and does not necessarily represent the views of the Council.

## Conseil International pour l'Exploration de la Mer

[^0]
## TABLE OF CONTENTS

0 EXECUTIVE SUMMARY ..... 1
0.1 ToR ..... 1
0.2 Assessments and state of stocks ..... 1
0.2.1 Pandalus in the Skagerrak and Norwegian Deep (IIIa \& IVa East) ..... 1
0.2.2 Pandalus in the Barents Sea and Svalbard area (Sub-areas I \& II) ..... 1
0.3 The by-catch in the Pandalus fisheries ..... 1
0.4 Working procedures ..... 1
1 TERMS OF REFERENCE. ..... 2
2 PARTICIPANTS ..... 2
2.1 The establishment of a joint NAFO-ICES Pandalus WG. ..... 2
3 DEFINITION OF STOCK / ASSESSMENT UNITS ..... 3
3.1 The North Sea and Skagerrak .....  3
3.2 The Barents Sea and Svalbard area. ..... 3
4 THE PANDALUS STOCK IN DIVISIONS IVA EAST AND IIIA ..... 4
4.1 The Pandalus fisheries in the North Sea and Skagerrak ..... 4
4.1.1 The Danish Pandalus fishery .....  4
4.1.2 The Norwegian Pandalus fishery .....  4
4.1.3 The Swedish Pandalus fishery ..... 5
4.2 Landings, catch and effort data ..... 6
4.2.1 Landings .....  6
4.2.2 Discards ..... 6
4.2.3 Effort data ..... 6
4.3 Sampling of Landings .....  7
4.3.1 Sampling frequency, intensity .....  7
4.3.2 Catch in numbers at age ..... 7
4.3.3 Mean weights at age ..... 7
4.3.4 Estimation of SSB, maturity ogives ..... 7
4.4 Trawl survey data .....  8
4.4.1 The Norwegian trawl survey ..... 8
4.4.2 Analysis of Survey Data ..... 8
4.5 Assessment of the Pandalus stock in Divisons IIIa and IVa East ..... 9
4.5.1 Background / history ..... 9
4.5.2 Stock Production Model applied in the 2003 assessment ..... 10
4.5.3 State of stock in 2004 and 2005 ..... 12
4.5.4 Biological Reference Points. ..... 12
5 THE PANDALUS STOCK ON FLADEN GROUND (DIVISION IVA) ..... 13
5.1 Catch and Effort. ..... 13
5.2 Previous Assessments ..... 13
6 THE PANDALUS STOCK IN FARN DEEPS (DIVISION IVB) ..... 14
7 THE PANDALUS STOCK IN THE BARENTS SEA AND SVALBARD AREA ..... 14
7.1 Description of the fisheries and their development in recent years ..... 14
7.2 Landings ..... 14
7.3 Discards ..... 15
7.4 Effort and CPUE ..... 15
7.5 Sampling of landings ..... 15
7.6 Research Vessel Data. ..... 15
7.6.1 Trawl Surveys ..... 15
7.6.2 Analysis of Survey Data area. ..... 16
7.6.3 Natural mortality and predation ..... 17
7.7 Assessment of the Pandalus Stock in the Barents Sea ..... 17
7.7.1 Background ..... 17
7.7.2 Status of the Stock ..... 17
7.7.3 Recommendations on further work ..... 17
8 THE BY-CATCH IN THE PANDALUS FISHERIES IN THE SUBAREA IV AND DIVISION IIIA ..... 18
8.1 Available data ..... 18
8.2 The magnitude of cod landings from the Pandalus fisheries. ..... 18
8.3 Improved species selection in shrimp trawls equipped with selective grids. ..... 19
9 THE BY-CATCH IN THE PANDALUS FISHERIES IN THE BARENTS SEA ..... 19
10 REFERENCES ..... 20
ANNEX 1 - ON THE MINIMAL ALLOWABLE SIZE OF THE BARENT SEA NORTHERN SHRIMP ..... 59
ANNEX 2 - THE SPANISH NE ARCTIC SHRIMP FISHERY IN 2003 ..... 70
APPENDIX 1 - PARTICIPANTS LIST ..... 74

### 0.1 ToR

The ICES Pandalus Assessment Working Group [WGPAND] met in ICES HQ, 27 Oct. - 4 Nov. Five participants from four countries attended the meeting, representing Denmark, Norway, Sweden and Russia. The terms of reference for the Working group were: to carry out assessments of the stock of Pandalus in Skagerrak and the Norwegian Deeps (IIIa \& Iva East) and the stock in the Barents Sea and Svalbard area (I \& II) and to provide advice (catch options) for these stocks in 2005.

### 0.2 Assessments and state of stocks

### 0.2.1 Pandalus in the Skagerrak and Norwegian Deep (IIIa \& IVa East)

In recent years the assessment of this stock has been based on a simple production model taking predation into account. The main input data to this model has been provided by a Norwegian trawl survey. However, because there has been a significant break in the data series in 2003, no update assessment could made this year due to missing assessment data. Furthermore, it has also been pointed out, that the model applied has not been optimal to the data available, see Sects $4.4 \& 4.5 .1$.

The state of this Pandalus stock in 2004 and 2005 is presented in Sect. 4.5.3. It is based on WG considerations on trends in LPUE combined with the 2004 biomass index provided by a new Norwegian survey. The stock seems to be on the same, rather high, level as in recent years

### 0.2.2 Pandalus in the Barents Sea and Svalbard area (Sub-areas I \& II)

Several models have been applied for assessment of this stock. However in recent years, the views on the state of the stock have been based on survey indices combined with trends in CPUE. SSB appears to have been declining in recent years, see Sect. 7.7.2.

### 0.3 The by-catch in the Pandalus fisheries

Sects. $8 \& 9$ give overviews of the by-catch based on mainly available logbook information

### 0.4 Working procedures

This years WG meeting was held jointly with NAFO Scientific Council/STACFIS shrimp assessment meeting at the same dates and at the same place (ICES HQ). The purpose of such joint meeting was to exchange views and experience in data and methodologies in assessments Pandalus stocks. However, it became clear, - and here was agreement among the NAFO and the ICES scientists -, that the scientific benefits were limited. This was mainly due to lack of crucial assessment input to the ICES WGPAND from one of the member countries. Thus, the basis for exchange and discussion of the assessments and methodology was severely constrained.

## The terms of reference are according to 2003 Council Resolution 2ACFMND

Pandalus Assessment Working Group [WGPAND] (Chair: S. Munch-Petersen, Denmark) will meet at ICES HQ, Copenhagen, Denmark from 27 October to 5 November 2004 to:

- assess the status of the stocks of Pandalus borealis in the North Sea, Skagerrak and Kattegat and provide catch options for 2005 taking predation mortality on Pandalus stocks into account;
- for stocks for which a full analytical assessment is presented, comment on this meeting's assessments compared to the last assessment of the same stock.
- assess the status of the shrimp stock (Pandalus borealis) in the Barents Sea, taking predation by cod into account;

WGPAND will meet jointly with NAFO Scientific Council/STACFIS shrimp assessment meeting at the same dates and at the same place. NAFO Scientific Council, STACFIS and ICES WGPAND chairs will jointly agree on the meeting arrangements. The arrangements will be made with a view of limiting the meeting time for WGPAND) and in particular ensure that the assessment of the Pandalus borealis stock in the North Sea, Skagerrak and Kattegat will take the abundance survey results into account.

## 2 Participants

| Aschan, Michaela | Norway |
| :--- | :--- |
| Berenboim, Boris | Russia |
| Eigaard, Ole | Denmark |
| Munch-Petersen, Sten (Chairman) | Denmark |
| Ulmestrand, Mats | Sweden |

In addition, Knut Sunnanå (Norway) contributed with data and suggestions by correspondence.

### 2.1 The establishment of a joint NAFO-ICES Pandalus WG

This year's ICES Pandalus Working Group meeting was held jointly with the NAFO STACFIS shrimp assessment WG. Both groups felt that joint meetings would be of great benefit to the work of the groups. The purpose of such joint meetings is to provide greater peer review of the assessment of Pandalus stocks. In addition such meetings will allow the exchange of information on assessment methods and on shrimp population dynamics. Unfortunately this first attempt at a joint meeting was less than successful, with the number of joint sessions being far less than had been planned. Part of this problem was lack of crucial assessment data as well as poor information to the WGPAND members on these data problems prior to the meeting. As a result the meeting developed into parallel sessions with only few joint scientific sessions and this is not the optimal way for joint meetings to be held. However it was felt by both groups that this problem could be overcome in the future and that the success of future meetings would be enhanced by having preliminary assessments ready for all stocks at the beginning of the meeting.

It was agreed that a joint NAFO-ICES Pandalus WG should meet in October 2005. In order to facilitate future meetings the Chairs of Scientific Council, STACFIS and WGPAND will work together to prepare a proposal for a plan for future meetings. This proposal will first be reviewed by members of the two groups, intersessionally and then
presented to Scientific Council at its June 2005 meeting. In devising such a plan the Chairs will consider the merits of running the STACFIS/WGPAND section of the meeting under a single chairmanship and they will also consider the possibility of incorporating review of methodological developments in future meetings.

## 3 Definition of Stock / Assessment units

### 3.1 The North Sea and Skagerrak

The distribution of Pandalus in the Entire North Sea area is shown in Fig. 3.1. The WG has, so far, maintained the view that shrimp caught on the Fladen constitute a stock separated from the Pandalus in the Norwegian Deeps and Skagerrak. The main arguments for this separation were presented in ICES (1990):

- Geographical separation combined with hydrographical considerations.
- The Fladen shrimp are normally characterized by fewer age groups This difference was quantified by multivariate analyses of length frequency distributions (LFD) from the three areas, these suggested that especially the Fladen LFDs deviate from the other two (ICES, 1990).

A close connection between the shrimp in the two areas has, however, been postulated by earlier investigations (e.g. Poulsen, 1970). It was done based on trends in size distribution of the shrimp in various parts of the entire North Sea - Skagerrak area and on probable larval drift with surface currents in the northern North Sea. The WG has, furthermore, observed that:

- Norwegian Survey data on recruitment for IIIa ,IVa East and LPUE in the Danish Fladen fishery is correlated pattern in LPUE fluctuations in the fisheries exploiting the two stocks have frequently been similar

This could indicate a close connection between the two stock units.

Improvements in genetic separation technologies in recent years could elucidate this particular stock separation problem. It is recommended that data for genetic analyses should be provided from the Norwegian trawl survey.

### 3.2 The Barents Sea and Svalbard area

The Pandalus stock in the Barents Sea and Svalbard area is distributed as shown in Fig. 3.2. Genetic investigations, allozyme electrophoresis and DNA-fingerprinting, have been conducted in attempts to identify potential sub populations of shrimp in the Northeast-Atlantic including the Jan Mayen area, the Norwegian coast, the Barents Sea and the Svalbard area (Kartavtsev et al. 1991, Rasmussen et al. 1993, Drengstig et al. 2000 and Martinez et al. 1997). The latter analyses showed that there are no distinct sub-populations in the open sea, and that there is a high degree of genetic variance between individuals within each location. However, genetic gradients related to geographic distance and sea currents have been identified. The transport pattern produced by the currents, varies between years, and results in annually different dispersion patterns of settled shrimp larvae. This may have a strong influence on the year class strength in sub-areas as well as in the entire Barents Sea.

The shrimp in the Barents Sea should be considered as one population, where female shrimp produce settling larvae in the whole distribution area. The transport of larvae secures genetic Flow within the population. The abundance of reproducing females in each sub area is of great importance for the annual recruitment and therefore management has to secure the spawning females throughout the Barents Sea (Pedersen et al. 2003).

### 4.1 The Pandalus fisheries in the North Sea and Skagerrak

### 4.1.1 The Danish Pandalus fishery

The Danish Pandalus fishery targets both the shrimp stock in the Sub-area IVa East and division IIIa and the one on Fladen Ground. In the period 1994 to 1999 the fisheries in the two areas were of about the same size, but since 2000 the landings from IVa-East + IIIa have been much higher and today they are almost three times the size of the landings from Fladen Ground. Interview information from the fishing industry obtained in 2004 gives the explanation that this shift in recent years is caused by poor fishery on the Fladen Ground (low abundance of shrimp) combined with low prices on shrimp and high prices on fuel. These latter conditions have further favoured fishing in waters close to landing harbours (Skagerrak) in order to minimize fuel costs.

During recent years an increasing number of vessels have started processing (boiling) the shrimp aboard and landing them in Sweden thus obtaining a better price. The majority of the catches are however still landed in Danish fishing ports. Most shrimp are landed directly to a few large factories processing almost all sizes of shrimp.

## The fishing vessels

In a study performed by Ulrich and Andersen (2004) all Danish fishing vessel were grouped in categories based on similarities in catch composition, gear used and area fished. According to their analyses of log book data on catch, effort as well as landings from all the Danish fishing trips in 1999, a total of 14 vessels could be identified as being trawlers targeting Pandalus in the North Sea or Skagerrak. They accounted for the majority of the total landings and had an average of 68 yearly trips targeting Pandalus. A larger poorly defined vessel group occasionally took part in the Pandalus fishery, but only accounted for small catches of Pandalus.

This result is in agreement with the interview information where Skagen, Hirtshals and Hanstolm were pointed out as being the major harbours of Pandalus trawlers in 2004, Skagen being the home harbour of 6-7 vessels of approximately 100-200 GRT and Hirtshals and Hanstolm each having 2-3 somewhat larger Pandalus trawlers of between 200 and 300 GRT. The major landing harbours were the same. According to the interviewees the number of Pandalus trawlers had been slightly higher in previous years.

## Fishing Gear

The largest net manufacturer in Denmark (Cosmos Trawls) provides shrimp trawls to many of the Danish vessels. At present the two most common trawls are the "Sputnik" (or "Skagerrak") trawl and the "Fladen shrimp" trawl, differing mostly with respect to the height of their trawl opening. The Sputnik trawl has almost twice the height as that of the Fladen shrimp trawl but only a slightly larger width. The two trawls are chosen by turn depending on fishing area and time. The mesh size in the cod ends used is almost exclusively 40 mm whole-mesh with a 70 mm square mesh window in the top panel.

This net manufactory also informed that within the last 5-10 years almost all trawlers had started fishing with twin trawls. This change had allowed the individual vessels to increase the swept area (wing end to wing end) with approximately $50 \%$ without resulting in increased demands to the vessels engine capacity or in any noticeable increase in fuel consumption.

### 4.1.2 The Norwegian Pandalus fishery

In 2002, a total of 143 trawlers were registered in three categories of shrimp trawlers conducting shrimp fishery mainly south of $62^{\circ} \mathrm{N}$. There were 45 vessels being less that 50 GRT and smaller than 13 m in length delivering 980 t of shrimp from this area, there were 69 trawlers less than 50 GRT and longer than 13 m , delivering 2770 t of shrimp and finally, there were 29 trawlers being larger than 50 GRT delivering 2330 t of shrimp. Vessels belonging to other categories also land some shrimp. According to the Norwegian logbook records for 2003, 38 vessels have reported shrimp catches and these vessels are all longer than 13 m . Of the 18 vessels less than 50 GRT, 4 vessels deliver less than
$10 \mathrm{t}, 10$ vessels between 10 t and 50 t , and 4 vessels more than 50 t . Of the 20 vessels larger than 50 GRT, 2 delivered less than $10 \mathrm{t}, 3$ between 10 t and $50 \mathrm{t}, 5$ between 50 t and 100 t and 10 more than 100 t .

In the Norwegian fishery for shrimp in this area the minimum mesh is 35 mm . It is not allowed to fish in waters shallower than 60 m . It is allowed to have $50 \%$ by-catch of other market species. For cod and haddock combined there is a limitation that the number of undersized specimens may not exceed 8 per 10 kg of shrimp. It is allowed to have up to $10 \%$ undersized shrimp ( $<6 \mathrm{~cm}-15 \mathrm{~mm}$ carapace length) in the catch. Discarding is prohibited in the Norwegian waters. Due to these regulations, the trawlers fish a considerable by-catch of market fish. They also conduct other fisheries during the year, e.g. mackerel trolling. The larger vessels ( $>50$ GRT) also conduct trawl fishery for sandeel and herring.

In 1999 a general quota regulation system was initiated in the Norwegian fishery. The total Norwegian quota is divided into periods of four months each with app. $1 / 3$ of the quota each period. The quota for 2003 was set at 8040 t for the whole year. The vessels have a maximum quota each for each period, a trip-quota for each trip to sea and a mandatory number of days of no fishing between each trip. There is some variation depending on whether they are fishing for boiled landings or for shrimp to be landed fresh.

Two categories of shrimp dominate the market: Approximately $35 \%$ of the total landings is delivered as boiled or fresh large shrimp (140-150 individuals per kg) for the Norwegian inland market (app. 60\%) and the Swedish market (app. $40 \%$ ) and app $65 \%$ of the total as raw (smaller) shrimp for factory processing (mostly 180-250 individuals per kg). A price and quota regulation is in work to regulate the available shrimp for the Swedish market, for which there is an export quota free of toll. The fisher gets app. 55 NOK for boiled shrimp and app. 10 NOK for the raw shrimp. Some high grading and discarding is assumed to take place. Especially shrimp sized below 15 mm carapace length are probably all discarded and may account for $5-10 \%$ of the catches (app. $6 \%$ in October 2003 based on the catches in the Norwegian survey).

### 4.1.3 The Swedish Pandalus fishery

In 2003, a total of 75 trawlers reported landings of Pandalus in the Swedish log books. Of these 52 landed more than 10 tonnes Pandalus and can be considered active in this fishery. 15 of these landed more than 50 tonnes and are specialized Pandalus trawlers, standing for about $50 \%$ of total Swedish Pandalus landings.

The size of the vessels ranges between $8-34 \mathrm{~m}$ (length) with an average of 18.7 m . GRT varies from 3 to 235 ), with an average of 81 GRT. The average engine effect is around $300 \mathrm{~kW}(44 \mathrm{~kW}-735 \mathrm{~kW})$. The larger trawlers are normally fishing in the eastern and central part of Skagerrak. The smaller trawlers are mostly fishing in the Swedish coastal zone inside a 'trawling border' where special regulations apply for the use of trawls: Trawling is restricted to waters deeper than 60 m and there are special limits in the length of ground rope and in the size of the trawl and trawl doors. Furthermore, the trawls to be used inside this boarder must be equipped with a species selective Nordmøre grid of 19 mm bar space and an unblocked fish opening in the trawl roof. This has resulted in very clean landings from these trawls ( $99 \%$ Pandalus). The Nordmøre grid may also be used outside the trawling boarder as an alternative to the EU legislated 70 mm square mesh panel in shrimp trawls.

This particular Pandalus trawl can be distinguished from other shrimp trawls in the log books since 1997 and it seems that the effort of this gear has stabilized at about $12 \%$ of total Pandalus trawl effort in the recent two years, see also Table 8.1. There are two different Swedish markets for Pandalus: a) higher value boiled larger sized shrimp constituting around $50 \%$ of the landings b) lower value smaller sized shrimp landed fresh and sold to the industry for further processing. The boiled Pandalus landings are cooked onboard before landed. Since the shrimp loses weight when boiled, these landings must be raised by a factor of 1.13 to obtain fresh weight for the landings statistics.

The TACs are limiting the Swedish Pandalus fishery and in order to distribute landings over the year the fishers have voluntarily introduced rations per fisher per week. This has resulted in high-grading of the catch, increasing the discarding of less valuable smaller Pandalus to increase the proportion of the more valuable boiled shrimp in the individual landings ration.

### 4.2.1 Landings

Landings are given in Table 4.1 by area (Division IIIa and Sub-area IV) as officially reported to ICES. In Skagerrak the landings for 2003 increased approximately $10 \%$ compared to 2002. It was only in the Danish fishery the recorded landings increased. In Sub-area IV total landings have increased only slightly, due to increased Norwegian landings in 2003. The combined total landings from IIIa and IV were 8\% higher in 2003 than in 2002.

Table 4.2 presents the landings and estimated discards for the assessment unit 'Skagerrak and the Norwegian Deeps' i.e. Div. IIIa and the eastern part of Div. IVa. The landings in 2003 were around 13000 t , an increase of 1000 t compared to landings in 2002. Some errors and misprints in the landing figures for earlier years were corrected.

Landings from Norway and Sweden (and to a very small extent from Denmark) consist of a fraction of larger shrimp that are boiled on board and a remaining portion of smaller shrimp landed fresh. The boiling causes the shrimp to loose weight. The conversion factor to obtain live weight is 1.15 . Official reported figures from Norway are given as landed weight. Sweden has adopted the same procedure for the last few years. In the amounts used by Working Group, the Swedish landings of large shrimp have, however, always been converted to live weight. The amount added for 2003 was 145 tonnes. The Working Group has applied no conversion on the Norwegian landings. The underestimate of total landings by this omission was for 2000 roughly estimated to about 300 ton. The Working Group felt that this estimate was too inaccurate to include in the assessment figures. When more reliable data for estimations become available, the landings for all years should be updated.

### 4.2.2 Discards

In the Norwegian and Swedish fisheries one may distinguish two categories of discarded shrimp:
The smallest size fractions from the grading procedure are not accepted by the canning industry and are discarded. This practice is traditional in the Norwegian and Swedish fisheries. This is probably also the case for the Danish catches. The Working Group estimated the amounts of discards by using the Norwegian length measurements from samples taken onboard before discarding. The proportions below $15-\mathrm{mm}$ carapace length are considered to be discarded. The estimated amount for 2002 was 254 t . Since these Norwegian data were not available for 2003, no such estimate is presented for 2003. However, it is likely that the amount of discards has been much higher in recent years than suggested in Table 4.2.2.

Quota restrictions and the substantial price difference between large, boiled shrimp and medium sized fresh ones together with a voluntary system of weekly rations (different for medium and large shrimp) have resulted in high grading by discarding the medium sized ones. In recent years several Danish shrimp vessels landing boiled shrimp in Sweden have probably been following this practice. The amounts of discards in this category in the Norwegian and Swedish fisheries were in an earlier report estimated for 1996 and 1997 only. The estimation was based on separate quarterly length distributions for the categories large and medium sized and the selection ogive for the sieve. The total annual amount of this type of discards could be more than 1000 tons. However, at present such estimates are considered too inaccurate to be included in assessments, but the working group expects that better data on discards will available through the current EU funded discard sampling programmes. According to qualitative information from the Danish fishing industry, the amounts of discarded shrimp in the Danish Pandalus fishery are rather small.

### 4.2.3 Effort data

Annual national figures for landings per unit of effort (LPUE) and estimated effort are shown in Table 4.3 and Fig. 4.1. Total national effort values have been estimated from LPUE data based on logbook records. The Danish and Swedish LPUE in 2003 was on the same level as in 2002, whereas the Norwegian LPUE seems to have increased. However, according to information from the Danish fishing industry the majority of the Danish shrimp trawlers have been equipped with twin trawls during the last 10 years period increasing the efficiency of vessels. The EU logbooks do not give information on the number of trawls used, and quantitative information on the development in the Danish fishery is not available yet, but it is likely that the trend of the recorded Danish LPUE figures during the last 10 years is biased to some extent, and that the real LPUE is overestimated and therefore the total effort may be underestimated. It is recommended that the qualitative information on the development of the Danish shrimp trawls be quantified in order to obtain more realistic estimates of LPUEs. The Swedish shrimp trawls are still mainly single trawls. Here it was pointed out that the discarding due to high grading, the Swedish LPUE figures may be underestimates. No information on the development in the Norwegian shrimp gear for Skagerrak and the Norwegian Deeps was available.

In order to obtain the same effort unit for all 3 countries, i.e. 'fishing hours', the Danish unit 'fishing days' was transformed to 'hours' on basis of functional regressions between Danish-Norwegian and Danish-Swedish LPUE. These two regression coefficients were averaged to get Danish $\mathrm{kg} / \mathrm{hr}$ as well as the total Danish effort in hours (unit $=1000$ hours). The missing Norwegian data from 1984-85 were estimated by functional regression NorwaySweden and the factor 1.12 applied. The estimated time series of total international effort (Khrs) and LPUE ( $\mathrm{Kg} / \mathrm{hr}$ ) are shown in Table 4.4 and Fig. 4.2

### 4.3 Sampling of Landings

### 4.3.1 Sampling frequency, intensity

National sampling effort is presented in table 4.5. The overall sampling level 2003 was around 15 kg per 1000 ton landed or 2500 specimen. Variations in the intensities between countries and between seasons indicate that improvements could be made.

### 4.3.2 Catch in numbers at age

The length data are pooled by quarter, and these national quarterly length distributions have then been partitioned into age compositions by the Bhattacharya method (software: FISAT). As in previous years the mean lengths by age group are used as a check of the consistency of the estimates, see Figure 4.1. Due to lack of Norwegian length data for 2003 the Norwegian total landings were age distributed according to the combined Danish and Swedish age data.

Table 4.6 gives the "catch-at-age" data. While previous years' tables also tabulated landings at age, this year's tables have included discarded 0 and I-Group shrimp. Catches are dominated by shrimp of ages 1 and 2 . It is seen that in 2003 there seems to be shift towards younger ages in the stock. The numbers of age 3 and older are likely to be underestimates, due to the way the Bhattacharya method operates. In general, the WG doubts the reliability of estimates of the older age groups, i.e. those $>$ age 3 . This doubt is also reflected in the pooling of ages $>3$ in to a 'plus-group' in the XSAs performed in previous years.

### 4.3.3 Mean weights at age

Weights-at-age for the Danish catches were derived from the length samples of the catches, where the weights of the measured shrimp in each sample are recorded by length group. The corresponding Norwegian and Swedish weights-at-age figures are based on quarterly length-weight relationships obtained from the Swedish length samples in which all shrimp are weighted individually. The mean weights-at-age in the catch is given in Table 4.7. In some years there were no records 0 -group shrimp in the catches, then averages for the other years were used. The same procedure was applied for the + group (+gp) in 2003.

### 4.3.4 Estimation of SSB, maturity ogives

In the estimation of SSB for the Pandalus stocks in the North Sea area the 0 - and 1 -groups are assumed to be immature, and age group 3and older groups are fully mature. In the cohort based assessments (XSA) the mature part of the 2 -group or potential spawners was taken as the sum of intersexes and females in the first quarter of the year.

These proportions were:

| 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0.62 | 0.09 | 0.20 | 0.26 | 0.82 | 0.96 | 0.73 | 0.59 | 0.45 | 0.70 | 0.51 | 0.58 | 0.51 | 0.60 | 0.65 | 0.76 | 0.51 | 0.52 |

This method was not appropriate in connection with the SPP model, where stock size estimates were based on survey data from October-November. At the time of the survey it is assumed, that the spawning stock consists of the females in the stock. Thus to get estimates of SSB the proportion of females (by weight) in the survey catch was calculated. These proportions varied between 56 and $74 \%$. They were then applied on the estimates of average biomass, $\left(\mathrm{B}_{\mathrm{y}}+\mathrm{B}_{\mathrm{y}+1}\right) / 2$.

Since no assessment was made this year, no update of the maturity data has been presented. The WG recommends, that in future assessments the procedures for estimating SSB be standardised for all Pandalus stocks in the North Atlantic.

### 4.4 Trawl survey data

### 4.4.1 The Norwegian trawl survey

The Norwegian trawl survey for shrimp in the Skagerrak-Norwegian Deep area has been conducted during six weeks in October-November from 1984 to 2003. However in the two recent years, 2003 and 2004, significant changes have taken place. The R/V 'Michael Sars' used from 1984 to 2002 was taken out of service in 2003 and replaced by R/V 'Haakon Mosby' for the survey in 2003. This vessel, at that time, had winches that were too small to haul the standard Campelen survey trawl and a different trawl was used. In 2004 the R/V 'Haakon Mosby' had the winches from R/V 'Michael Sars' installed. The survey period was, however, shifted to May for several reasons - the most important was availability of vessel due to removal of R/V 'Michael Sars' from the fleet.

The survey area is divided into 16 strata: from 100-200m (stratum 1,3,5,8,11 and 14), 200-300m (stratum $2,4,6,9,12$ and 15) and $300-500 \mathrm{~m}$ (stratum 7,10,13 and 16), see Fig. 4.4. The trawl hauls are spread evenly over the area covering the depths mainly between 100 m and 450 m depth. On the southern and eastern part of the Deeps it is possible to trawl almost everywhere, whereas along the Norwegian coast the areas suitable for trawling are restricted. For most of the trawl hauls the same positions are used all years. The gear used, except for 2003, is a "Campelen 1800 \# x 35 mm " shrimp trawl, which is the standard demersal sampling trawl of Norwegian research vessels. A rubber-wheel gear was used until 1989 when a rockhopper gear was mounted on the same trawl. The opening of the trawl is app. 4.5 m and the wingspread is app. 15 m . During shrimp surveys a 6 mm mesh-size lining net in the cod end is used whereas the meshes in the rest of the cod-end is 22 mm .

In 2003 an $\AA$ Åkra Shrimp trawl 1420\# x 40mm was used, having a 7.5 m opening height and a wingspread of app. 20 m . The mesh size in the cod end is 42 mm , although an inner lining net was used.

The duration of the trawl hauls was 1 hour until 1989, since then 0.5 hour hauls have been the practice. It is assumed to be of great importance that the same vessel, fishing-gear and time of year be used in order to obtain the best abundance estimates. However, these will also be influenced by annual variation in temperature and other environmental conditions. Especially the 0 -group has through time proved to be unreliable as a measure of year class strength as temperature influence growth and vertical distribution.

In 2004 the survey was undertaken in May. However, it is strongly recommended that the future surveys be conducted in the 1 quarter in order get good estimates of the 1 group as recruits and good estimates of SSB (berried females). Together with the odd trawl in 2003 the time series is at present suffering a severe "hole" or "break" that may cause problems in the assessment, now and in the future. It is therefore recommended that calibration experiments are conducted to obtain correction factors for shrimp and fish (weight and length distributions).

In this connection the WG wishes to stress this year's problems with the lack of crucial Norwegian assessment data. When WG letters specifying data requirements have been sent well in advance to all WG members, one would expect response from all active members providing a major part of the necessary data input, especially when significant changes in time series have taken place since the last assessment meeting

### 4.4.2 Analysis of Survey Data

### 4.4.2.1 Shrimp: Swept area estimates of numbers-at-age

The width of the trawl opening, used for calculating swept area estimates, is 11.7 m . (Teigsmark \& Øynes, 1983). The average speed is $3 \mathrm{~nm} /$ hour and thus the trawl covers 0.019 nm 2 in 1 hour. The total number of shrimp caught per hour is calculated for each trawl haul. Total number of shrimp in a stratum is calculated as average number of shrimp per hour by stratum, divided by 0.019 and multiplied by the area of the stratum.

Samples of 250-300 specimens are taken from each trawl haul. The shrimp are measured (CL) to the nearest mm below. Weighted average length frequencies for each stratum are calculated with catches per hour in the trawl hauls used as weights. This length distribution is then split into age groups by the Bhattacharya method. The output is number
of shrimp by stratum and age group. These figures are considered as indices only, since a proportion of the shrimp is pelagic and not available to bottom trawl. This is especially the case for the 0 -group (Table 4.8).

### 4.4.2.2 Shrimp: Swept area estimates of total biomass

Total weight of shrimp caught per hour is calculated for each trawl haul. Total weight of shrimp in a stratum is calculated as average catch of shrimp per hour by stratum, divided by 0.019 and multiplied by the area of the stratum. The output is weight of shrimp "available to trawl" by stratum (Table 4.9).

In order to get an estimate of total biomass in the surveyed area, a value for the catch efficiency (= catchability coefficient) of the trawl is needed. The only "estimate" of the efficiency of the "Campelen 1800" trawl that is available is the 11.7 m "efficient" wing spread used in the swept area calculations. To obtain a better estimate of the total biomass the WG therefore has also applied, as a first approximation, an additional value of 0.2 for catch efficiency. This value is based on the work of Berenboim et al., 1980. These authors describe photogrammetric investigations in the Barents Sea where shrimp density (numbers per square meter) were measured. It was found that a 'catch efficiency' of 0.173 gave abundance estimates corresponding to the photo survey. Applying a catch efficiency of 0.2 on the estimated figures for biomass "available to trawl" given in Table 4.9 estimates of total biomass in the area are obtained (see Table 4.15).

### 4.4.2.3 Fish: Catches per nautical mile

The fish in the hauls were measured and weighed by species. Biomass by species and nautical mile trawled was calculated by haul and averaged over all hauls. This procedure is used, because the database is designed to provide fish data by statistical rectangles and not by the 16 strata as for shrimp. (Table 4.11.)

### 4.5 Assessment of the Pandalus stock in Divisons IIIa and IVa East.

### 4.5.1 Background / history

The Pandalus stock in Div. IIIa and IVa East was assessed by cohort analyses (VPA/XSA) from 1987 to 2000. The input was age disaggregated catch data (Tables $4.6 \& 4.7$ ). Commercial CPUE at age was used for tuning in the earlier assessments. Later stock indices at age from the Norwegian shrimp surveys were used.

The XSAs were characterised by non-favourable diagnostics (see Sect 4.6 in ICES 2002/ACFM:04) and consequently by poor estimates of fishing mortality (F) and stock size (N). Several features characteristic to the shrimp stocks reduce the applicability of the XSA:

- Few age groups in the stock.
- Large uncertainties in the ageing of especially the older age groups.
- A variable natural mortality exceeding the fishing mortality.

An additional cause could be that the assumption of the Fladen stock being independent of the IIIa stock is wrong, cf. Sect. 3. The available age disaggregated data seem not, however, to fit the XSA model very well.

A connected problem has been the treatment of natural mortality. Even if natural mortality is likely to be variable owing to the high suitability of shrimp as prey item for many fish species, it was assumed constant at 0.75 in most of the assessments until 1999.

The WG attempted in its assessment 1999-2001 to include predator dynamics by letting the M values vary according to the total SSB of roundfish in the North Sea (sum of cod, haddock, whiting and saithe), and /or the weight of likely predator species caught in the Norwegian shrimp surveys. In want of better estimates of the likely level of predation these indices were scaled to give an average M of 0.75 over the time span of the assessment. However, this treatment of the predation resulted only in minor changes in the perception of shrimp biomass dynamics (Figure 3.3 in ICES, 2000).

In 2000 and 2001 several of the main input components to the assessment were subject to critical discussion and revisions in order to improve the assessment. A new assessment approach was introduced by applying a stock production model including a predator component. In this model (described below) predator components have been
included in the assessment of Divisions IIIa and East stock. This model, which is described below, has been applied for the assessment of this stock, 2001-2003. However, due to the break in the time series of the Norwegian trawl survey in Skagerrak and the Norwegian Deeps in 2003 consistent input data was not available to the 2004 WG. Furthermore, the WG also has taken notice of the problems and criticism of the simple SPP model used:

- The real level of biomass cannot be determined by the model, but has to be set by some catchability factor. However, no such factor is available for the Campelen 1800 survey trawl.
- The age 1 index is proportional to the bulk of the commercial catch and a dominant part of the survey index, and is thereby not an independent index of recruitment.
- The use of this simple production model is not suited with the available data without having growth and mortality parameters for the stock.
- The model is not very sensitive to the input figures, which then must be realistic to avoid the model to end up with unrealistic 'local' minima in the least square estimates.

The WG decided to leave last years SPP assessment in the report as background documentation for elucidation of the problems and for future improvements of the assessment.

### 4.5.2 Stock Production Model applied in the 2003 assessment

### 4.5.2.1 Description

For the assessments 2001-2003 the Working Group applied a model based on aggregated data. The model has been described and exemplified by Stefánsson et al. (1994). It is a stock production model (referred to here as SPP), which includes the effect of yearly recruitment and predation. Similar models have been described and evaluated by Cadrin (2000.)

The model is expressed as:
$\mathrm{B}_{\mathrm{t}+1}=\mathrm{aB} \mathrm{B}_{\mathrm{t}}-\mathrm{C}_{\mathrm{t}}+\mathrm{bR} \mathrm{R}_{\mathrm{t}}-\mathrm{pD}_{\mathrm{t}}$
$\hat{U}_{t}=q\left(B_{t}+B_{t+1}\right) / 2$
Where
$B_{t}-$ (fishable) biomass of shrimp in year $t$
$\mathrm{C}_{\mathrm{t}}$ - yearly catches
$R_{t}$ - yearly recruitment indices
$D_{t}$ - yearly index of predator biomass
$\mathrm{U}_{\mathrm{t}}$ and $\hat{\mathrm{U}}_{\mathrm{t}}$ are the reported and predicted biomass indices.
The parameters of the model $\left(\mathrm{B}_{0}, \mathrm{a}, \mathrm{b}, \mathrm{p}\right.$ and q$)$ are estimated by a least-squares fit (performed by the Excel Solver). The expression to minimise is the sum of squared differences between observed and predicted biomass indices:
$\Sigma\left(\mathrm{U}_{\mathrm{t}}-\hat{\mathrm{U}}_{\mathrm{t}}\right)^{2}$
The model was modified by changing the way biomass indices were calculated. In the original model yearly commercial CPUE were used as indices, these were related to the average biomass at the beginning and end of each year. The survey indices used here are related to the beginning of the year, thus:

$$
\hat{\mathrm{U}}_{\mathrm{t}}=\mathrm{qB}_{\mathrm{t}}
$$

### 4.5.2.2 Input Data for the 2003 assessment

The input data cover the period 1984 to 2002 and consisted of:
Yearly catches for this stock, see Table 4.12 and cf. Table 4.2.
Biomass indices from the Norwegian bottom trawl surveys in October raised to total area surveyed and expressed as thousand tonnes. Data points were moved forwards from November to 1 January. Strata 1,3, and 8 were omitted due to sporadic coverage by trawl hauls. From these total biomass values the weight of the 0 and 1 -group were subtracted, see Table 4.10.

Recruitment indices. The weight in ton (numbers * average weight) of 1-group shrimp raised to total area surveyed in the Norwegian Surveys. The number of 1-group shrimp was estimated by the Bhattacharya method (Table 4.10).

Indices of predation. The total weight of likely predator species ( 22 species) caught per nautical mile in the Norwegian trawl surveys in Skagerrak and NE North Sea was chosen. The dominating species in this assembly are Blue whiting ( $\geq 30 \mathrm{~cm}$ ), Saithe, Cod and Roundnose Grenadier. The species included were those that are known to consume Pandalus (Albert, 1993, 1994a,b, Bergstad 1991, Skjaeraasen 1998, Torres, Roderiguez-Marin and Loureiro 2000) (Table 4.11).

Note that the indices here refer to the estimated biomass "available to the trawl" in the survey. The survey estimates of 0 -group in the $4^{\text {th }}$ quarter were not considered useful as recruitment indices because of the varying ability of the trawl to catch them. The size ( $7-12 \mathrm{~mm} \mathrm{CL}$ ) of 0 -group shrimp is normally within the selection range of the gear, but both size and vertical distribution are depending on the temperature and vary between years.

Catches, recruitment indices and predator indices are presented in Table 4.12.

### 4.5.2.3 Results (2003 assessment)

Estimates of parameters values and the corresponding Sum of Squared differences (SSE) are given below.
a 0.99
b 5695
p 667
q $\quad 0.00012$
$\mathrm{B}_{0} \quad 90000$
SSE 195

Estimated shrimp biomass and amount consumed by predators are given in Table 4.13 together with the observed and calculated biomass indices.

The relation between observed and estimated biomass indices is also illustrated in Figure 4.5. The regression indicates that $42 \%$ of the variation in biomass indices could be explained by the model.

Exploitation expressed as Yield/Biomass has fluctuated between 8,6 and 19,9 \% with an average of 13,9\% (Table 4.14). Estimated predation mortality (Consumption/Biomass) has varied between $16,8 \%$ and $46,6 \%$ with an average of $30,7 \%$.

### 4.5.2.4 Establishment of biomass level (2003 assessment)

The model is not very sensitive for the chosen starting value of $\mathrm{B}_{0}$ (Stefánsson et al. 1994, ICES 2001). This was illustrated in Fig. 4.3 in the Working Group report from 2001 (ICES, 2001), which gives the SSE as a function of the $\mathrm{B}_{0}$ value.

The actual level of $\mathrm{B}_{0}$ was chosen on basis of a comparison with the estimate of total biomass in the Surveys using swept area and the chosen value of catch efficiency (ICES 2003). The Survey estimates and biomass estimated from the model with $\mathrm{B}_{0} 90000$ ton is given in Table 4.15.

### 4.5.2.5 Evaluation (2003 assessment)

### 4.5.2.5.1 Comparisons with Model Output and Commercial Fleet Data

## LPUE - Biomass Indices

The estimated international LPUE values were compared with biomass estimates from the SPP and with the survey indices (Fig. 4.6). For the early years of the period the correlation between LPUE and model biomass is not very convincing. The correlation coefficient is 0.5 .

## Effort - Y/B

A comparison between estimated international effort and $\mathrm{Y} / \mathrm{B}$ (as a proxy for fishing mortality) is shown in Fig. 4.7. Here $B$ is the 'average' biomass', $\left(B_{y}+B_{y+1}\right) / 2$. Both series give the same general trend; the correlation coefficient is 0.53 .

### 4.5.3 State of stock in 2004 and 2005.

No update assessment was made this year. However, the WG considered the available, most recent, indicators of stock biomass:

1) The trend in commercial LPUEs presented in tables $4.3 \& 4.4$ And Figs $4.1 \& 4.2$ as indicator of the development of the stock up to 2004: The combined LPUEs (Fig. 4.2) show an increasing long term trend from 1989 to a peak in 1997-8, declined again in 1999 to 2001 and increased in the recent two years and LPUE in 2003 is the highest observed during the period of available data. The combined effort shows a decreasing long term trend. The similar pattern in LPUE between the three countries might indicate that the exploitable stock biomass has followed the same pattern, i.e. increased until 1998, decreased in 1999 to 2001 and increased again in recent two years. The recorded increases in LPUE may also reflect increased fishing power of the shrimp vessels as stated in Sect. 4.2.3, where it is suggested that a shift to twin trawls could explain some of the increase in Danish LPUE.
2) The biomass index provided by the Norwegian survey in May 2004 compared to the indices up to 2002: Even if the swept area estimate of shrimp biomass for 2004 (Table 4.9) is not directly comparable with the estimates up to 2002, because of change of survey time, the 2004 figure certainly confirms that the stock seems to be stable.

Because the Norwegian survey index for 2004 indicates a stable stock it is the view of the WG, that the above mentioned trends in LPUEs (as biomass indicators) also reflect a stable stock.

From these considerations the WG concludes, that the Pandalus stock in the IIIa and IVa East area does not show signs of overexploitation. Also based on these considerations it is the opinion of the WG, that the stock is stable and will remain at the same level in 2004 and 2005 as in recent years.

On this basis the WG recommends, that the total landings from IIIa and IVa East in the coming years be maintained at a status quo level of around 15000 t .

### 4.5.4 Biological Reference Points

The view of the WG is that, the data on the stock-recruitment relationship, from previous assessments, did not support establishment of a SSB reference value for this Pandalus stock based on this relationship (ICES 2003). In 1998

ICES (ACFM, 1998) pointed out that there was not basis for establishment of a $\mathbf{B}_{\mathrm{lim}}$ on basis of the available S-R data at that time. Considering the major impact from predation, such a poor relationship is likely.

According to previous assessments, predation accounts for at least twice as much removal of the Pandalus stock than the fishery in the entire assessment period. Such dynamics also render it problematic to establish a reference value for F ( or $\mathrm{Y} / \mathrm{B}$ ), at least if the relative magnitudes of F and M (predation) are independent of stock size.

It is the opinion of the WG, that until progress is made in elucidating the SSB-R relationship as well as the F-M relationship, the best and simplest approach would be so set SSB reference values based on the some lowest "observed" (estimated) values of SSB. However, in the case of the Pandalus stocks in the North Sea areas such limit reference values, for instance as $\mathbf{B}_{\mathrm{pa}}$, could only be used in harvest control schemes when more is known on the predation. Until then they would serve as limit values triggering warning signals. At present, the magnitude of the stock is uncertain, but in this connection the WG notes that for data poor situations, where survey indices are applied as stock indicators it has been suggested that a decline of $85 \%$ from the highest observed index could be used as a proxy for $\mathbf{B}_{\text {lim }}$ (NAFO SCS Doc. 04/12).

The WG also notes that on Fladen Ground the Pandalus stock was seemingly depleted in 1988 after very high catches in 1987 (ICES, 1990). However already in 1989 the stock had recovered without any management regulations of the fisheries. This indicates that for shortlived species like Pandalus other factors than fishery must have major influence on stock development.

## 5 The Pandalus Stock on Fladen Ground (Division IVa)

### 5.1 Catch and Effort

Table 5.1 shows the landings from the Fladen Ground since 1972. Since 1991 total landings have fluctuated between a low of around 500 tonnes to a high of more than 5000 tonnes. Mainly the Danish and Scottish fisheries exploit the shrimp on Fladen. Denmark accounts for the majority of landings. Note that since 1999 total Fladen landings, like landings in IIIa and IVa East, have been on a lower level than in the previous 4 years period. The recent landings have only been around $1 / 3$ of the level of 1995-98 landings. Landings in 2002 declined further compared to the landings in 2001. No U.K (Scottish) fisheries on Fladen Ground directed for Pandalus took place in 2001, 2002 or 2003. In general, the shrimp fisheries on Fladen take place mainly during the first half of the year, mainly in the second quarter.

Total effort for the Danish and Scottish Fladen fisheries is estimated from logbook based LPUE data from these fisheries. Table 5.2 gives these effort data. In 2003 the Danish LPUE was at the same level as in 2002.Estimated total Danish effort decreased slightly due to a decline in landings. No effort data for 2001, 2002 and 2003 were reported from U.K. In order to combine Danish and Scottish effort data in the earlier years, relative effort indices have been calculated for each country. The combined effort indices have been weighted with the landings. Both LPUE and total effort has been at rather low levels in recent years compared to the higher levels in 1995, 19961998.

### 5.2 Previous Assessments

The shrimp stock on Fladen has not been assessed since 1992, due to incomplete age data and the lack of separate, fishery independent data. Thus the most recent analytical assessment of this stock was presented in the 1992 Working Group Report (ICES, 1992).

However, some data for analytical assessments for later years have been compiled at the national laboratories (Denmark and Scotland) and are available to the Working Group. However, due to the frequent large fluctuations in the Fladen fishery, samples for length composition of the catches do not always cover the entire year.

Catches from Fladen consist mainly of two age groups. During the first two quarters of the year age groups 2 and 3 normally dominate the catches. During the $4^{\text {th }}$ quarter age group 3 usually disappears from the catches, while age group

1 adds to the catches. Lack of information on recruitment from surveys in this area has prevented the Working Group from making stock predictions for the Fladen.

## 6 The Pandalus Stock in Farn Deeps (Division IVb)

Sine 1991, only UK vessels have fished Pandalus in the Farn Deeps. Total landings fell from 500 t in 1988 to none in 1993. In 1995 and 1996 again about 100 ton were reported. There have been no reported landings from the Farn Deeps in recent years, see Table 6.1.

## 7 The Pandalus Stock in the Barents Sea and Svalbard area

### 7.1 Description of the fisheries and their development in recent years

Norwegian vessels began to exploit the shrimp fisheries in the Barents Sea and Svalbard area in 1970. Russian vessels entered the shrimp fishery in 1974. The catches increased continuously until 1984 when the total catch reached a maximum of $128,000 \mathrm{t}$. By that time vessels from other countries had entered the fishery. Since then, biomass and catch levels have fluctuated because there were different recruitments, cod consumption and effort in the fisheries due to price of shrimp. The catch peaked above 80,000 t in 1990 and in 2000. The most important fishing ground is the Hopen area in the central Barents Sea.

The first vessels using double trawls entered the fishery in 1996. Since then the effort has increased continuously and in 2002 approximately 35 Norwegian vessels had the technology to use double trawl or even triple trawl. Since 2002 the majority of the catch is caught by double trawl.

In the Svalbard area the shrimp fisheries are regulated by number of effective fishing days and number of vessels by country. In the Barents Sea and Svalbard area, Norwegian rules are that the fisheries be regulated by fishing licences and by smallest allowable shrimp size (maximum $10 \%$ of catch weight may be $<15 \mathrm{~mm}$ carapace length, CL). However, the regulation by smallest allowable shrimp size is not considered to be an efficient management tool in the REZ (Bakanev and Berenboim WD1, Annex 1). In the Russian Economic Zone, a TAC is established each year by Russian authorities. Fishing grounds are closed if by-catch limits given as number of individuals in 10 kg of shrimp are exceeded. In 2004 and 2005 the values of allowed by-catch are set at eight for the sum of cod and haddock, ten for redfish and three for Greenland halibut per catch of 10 kg shrimp.

Sorting grids in the shrimp trawls first became mandatory operating within the Norwegian 12 miles zone in February 1990. In October 1991 this directive was extended to apply to shrimp trawls used in all of the Norwegian EEZ. Finally, in 1993 the Joint Norwegian Russian Fisheries Commission agreed that the sorting grid was to be mandatory for all vessels conducting shrimp fishery in the Barents Sea and the Svalbard area.

### 7.2 Landings

Preliminary reported landings for all countries show a substantial decrease of landings from 82,816t in 2000 to approximately $60,000 t$ in 2002 and 2001 and a further decrease to 40,000 t in 2003 (Table 7.1, Figure 7.1). Thereby the total landings have decreased to $50 \%$ in three years.

Since there is no TAC in the Barents Sea all catches are landed and it is believed that there are no discards of shrimp in the area. However, it is recommended that data on discards of small shrimp in the Barents Sea and Svalbard will be presented in 2005 .

### 7.4 Effort and CPUE

Catch, effort, and annual CPUE series for Norway and Russia are presented in Table 7.2. The CPUE series for both countries are given in Figure 7.2. The Norwegian shrimp fleet has since late 1990s been upgraded both concerning vessels and the use of double and triple trawls. In the logbooks the use of these trawl types have been difficult to register and to make available for further use. Here revised series of catch per unit of effort (new CPUE) and new effort have been given for Norway, standardised as vessels $1000-1500 \mathrm{hp}$ with single trawls. However, this standardisation does not seem to correct for all advancement in shrimp trawl technology. The Norwegian data show a peak in the effort in 2000 at the same level as the earlier peaks in 1985 and 1990. The Russian and Norwegian effort decreased in 2001 with a slight increase in 2002 followed by a further decrease in 2003. The CPUE of the Russian fleet (vessels<1300hp) has fluctuated in accordance with the shrimp biomass (Berenboim et al. 2001). It should be noted that the Russian fleet is also under development.

### 7.5 Sampling of landings

In 2002 and 2003 observers collected samples on board commercial Spanish vessels in the Svalbard zone (Casas Annex 2 WD 2). Length and sex distribution data and data on by-catch of young fish were recorded. These data show a reduction of females from $33 \%$ in 2002 to $18 \%$ in 2003.

Monitoring of the shrimp catches is required due to the regulation protecting juvenile fish and shrimp through area closures. The Directorate of Fisheries in Norway has, during surveillance cruises conducted by commercial shrimp trawlers, collected data on length distributions in the shrimp catch since 1995. The Norwegian Coast Guard also samples some length data during inspections of shrimp catches. In 2002 the Institute of Marine Research established a reference fleet where fishermen take samples of the catch. One of the vessels included in the reference fleet is a part time shrimp trawler. The carapace length is measured on 300 individuals of shrimp in each sample. The number of samples collected has increased from 70 samples in 1999 to 210 samples in 2003. The sampling frequency will be further increased by more inspections conducted by the Coast Guard.

The catch was in 2000 dominated by shrimp aged four and five years (Figure 7.3). The catch pattern moved towards three year olds in 2001. The catches in 2003 were again dominated by four year old shrimp of the 1999 year class.

### 7.6 Research Vessel Data

### 7.6.1 Trawl Surveys

In the Barents Sea and the Svalbard area, standard shrimp surveys have been conducted by Norway since 1982 and by Russia since 1984 (Figure 7.4.). However, during the 90 's, both surveys have suffered from reductions in survey time. The Russian vessels did not survey the Svalbard area for many years but have carried out surveys in this area in 2001 and 2002. Unfortunately no shrimp cruise was conducted by Russia in 2003 and 2004. The amount of time available for the Norwegian survey has been reduced from 50 days to 27 days. Detailed information pertaining to the status of the stock is described in 1981-1991 Norwegian reports (Tavares and Øynes 1980, Teigsmark and Øynes 1981, 1982, 1983a, 1983b, Hylen et al. 1984, Tveranger and Øynes 1985, Hylen and Øynes 1986, Hylen et al. 1987, Hylen and Øynes 1988, Hylen et al. 1989, Hylen and Ågotnes 1990) and Russian reports (Berenboim et al. 1986, Berenboim et al. 1989, Berenboim et al. 1990, Mukhin and Sheveleva 1991). Annual joint Norwegian-Russian papers have been produced since 1991 (Berenboim et al. 1992, Aschan et al. 1993,1994, 1995, 1996). Since 1997 the status of the stock has been summarised in annual protocols of the Russian-Norwegian Comission and ICES (Anon. 2003). Additionally evaluations of the Norwegian surveys have been conducted (Aschan and Sunnanå 1997, Harbitz et al. 1998).

The Norwegian shrimp cruises are conducted with R/V "Jan Mayen" in April-May in the Barents Sea and in August-September in the Svalbard area. In both areas more than 100 stations were sampled. In the Barents Sea a regular
grid placing the stations on a distance of 20 or 28 nautical miles, depending on shrimp density is used. In the Savlbard area the stations are also fixed but distributed within strata defined according to depth and latitude. The survey trawl is a Campelen with 1800 meshes and has the following specifications; 30 m headline, 19 m ground rope and $80-42 \mathrm{~mm}$ knot to knot stretched mesh size in the body. The mesh size of the cod end has been stable with 42 mm knot to knot stretched mesh. In 1986 a 4 m inner net of 10 mm stretched mesh size was introduced, and this was replaced in 1994 with a 15 m inner net with 20 mm mesh. A small mesh-bag ( 8 mm ) has been attached to the lower belly as standard equipment since 1995. The sample of small shrimp obtained in this bag gives an additional index of abundance of the smallest shrimp and provides data on the size of the youngest age groups.

Bridle arrangements have been constant at 40 m upper and lower bridles. A 24 m rope connecting the warps 80 m above the doors ("strapping") constrains the spread of the doors to 47 m , wing spread 14,5 meters and the vertical opening of the trawl to approx. 5 m ; the chains are equipped with two steel bobbins ( 45.7 cm diameter) each, 40 floats ( 20 cm diameter) are attached to the fishing line, 15 floats are attached on each side along the belly, and the mid sections of the "Rockhopper" gear are changed from metal to rubber discs.

Three types of trawl doors have been used. In the 1980s Waco combination doors ( 1500 kg ) were used. During the early 1990s traditional "Steinshamn" V10-doors ( 2050 kg ) were used, and in 1994 these were replaced by "Steinshamn" W9-doors ( 2050 kg ) with four point connections (Kristjansson, 1994). Thyborøn doors will probably be used in 2005 after an evaluation of results from parallel trawling.

In all Russian surveys, a commercial Russian shrimp trawl without sorting grid was used. The mesh size was 40 mm and a10m inner net with small-mesh $(12 \mathrm{~mm})$ was used. The horizontal opening of a trawl is 14.5 m , the vertical opening is 5 m . A trawling distance of 3 nautical miles was used and the trawling speed was 2.6-2.7 knots.

### 7.6.2 Analysis of Survey Data area

There is a strong correlation between the Norwegian and the Russian survey results (Figure 7.5). Biomass indices were highest during 1984, and have since fluctuated between $30 \%$ and $60 \%$ of this level (Tables 7.3. and 7.4.) with peaks in 1991 and 1998-1990 and low values below the long term mean in 1987-1988, 1994-1995 and 2001 to 2004. Norwegian bottom trawl surveys indicate a decrease in shrimp biomass in the Barents Sea and Svalbard of 29\% from 2003 to 2004. Especially the important Hopen Deep ( $-50 \%$ ) and the Thor Iversen Bank ( $-36 \%$ ) area show an obvious reduction.

## Swept area estimates of numbers-at-age

In order to obtain good length frequency distributions for age analyses, oblique carapace lengths (CL) (from the posterior margin of eyestalk to the posterior mid-dorsal edge) of approximately 300 individuals from each trawl station are measured to the nearest 0.01 mm with an electronic calliper (Mitutoyo, Japan). The data are saved in the database in intervals of 0.1 mm . Shrimp ageing is completed by modal analysis using MIX 3.0 (MacDonald and Pitcher 1979). Annual age determinations have been conducted for 15 areas in the Barents Sea and 7 areas in the Svalbard area since 1991 (Aschan 2001, Hansen and Aschan 2001).

Since the growth of shrimp varies in time and space, it is difficult to decide on a good recruitment index. An agelength key constructed from the Norwegian Barents sea survey has been used to define the number of 1,2,3 and 4 year old shrimp as well as the number of shrimp of five years or more in the whole Barents Sea (Table 7.5). A common procedure for dividing shrimp into age groups has been agreed upon. Since very few shrimp $<15 \mathrm{~mm}$ CL are caught in the trawl, the index for one year old shrimp is based on the number of shrimp caught in the mesh bag attached to the underbelly of the survey trawl. The recruitment index for one year old shrimp is still low and the number of two and three year old shrimp has reduced dramatically since 2003 (Table 7.5, Figure 7.6).

The biological development of shrimp is divided into several stages. Shrimp starts off as males (Stage 2 ) after the juvenile stage (Stage 1). Thereafter they reach intersex (Stage 3) before they develop into first time spawning females with headroe (Stage 4). When the females mate, the roe is moved under the abdomen (Stage 5) where the eggs stay until hatched (Stage 6). Some females then take a resting period (Stage 7), but the majority starts on a new cycle with headroe (Stage 8). The Russian and the Norwegian coding of the stages are given in Aschan et al. (1993). Analyses of data from the 90 's suggest that shrimp in the southern Barents Sea (area A) grew quickly and changed sex at an age of four years, whereas shrimp in the central and northern Barents Sea grew slowly (areas B, C and E) and changed sex at an age of 5 years or greater (Aschan 2001). The number of egg producing females is calculated using the annual $L_{50}$ in the Barents Sea. The number of egg carrying females (stage 5+), has declined since 2002 (Table 7.5, Figure 7.7).

### 7.6.3 Natural mortality and predation

Predation by cod is the main source of natural mortality. However, it should be noted that other fish species such as Greenland halibut (Reinhardtius hippoglossoides), long rough dab (Hippoglossoides platessoides), thorny skate (Raja radiata) and blue whiting (Micromesistius poutassou) also prey on shrimp (Dolgov 1997, Dolgova and Dolgov 1997. The methods used in estimating cod consumption are described by Bogstad and Mehl (1997), and dos Santos and Jobling (1995). In the Barents Sea, the annual consumption of shrimp was estimated to be above 280,000 t throughout the period 1994-2001 (Table 7.6, Figure 7.8). Shrimp consumption rates may, however, have been overestimated. As future shrimp assessments have to include cod as predator, it is important to identify and study possible problems with the cod consumption estimates.

### 7.7 Assessment of the Pandalus Stock in the Barents Sea

### 7.7.1 Background

The great plasticity in growth of shrimp and age at sex change, as well as a lack of biological data and length distributions from the catches, make it difficult to apply traditional analytical fishery assessment methods to the data. Therefore a spreadsheet performance report (Caddy 1999, Koeller et al. 2001) has been used to assess the available information (Table 7.7). Several models have been used in assessing shrimp in the Barents Sea and some of these are listed below: 1. Production models: Shaefer and Fox stock models and stock production model including predation (Stefánsson et al. 1994, Berenboim and Korzhev 1997). 2. Catch at age analysis (cohort models): Single species virtual population analysis (VPA) and multi species virtual population analysis (Sparre 1984, Bulgakova et al. 1995)).

### 7.7.2 Status of the Stock

The Russian CPUE (Table 7.2, Figure 7.2) and Norwegian survey indices (Table 7.3) both indicate a decrease in the shrimp stock from 2003 to 2004. The survey index of 2004 shows a reduction of $29 \%$ since 2003, and is now on the lowest level since 1987. The spawning stock number has been decreasing since 2002 (Figure 7.7). The recruitment of one year old shrimp has been low but stable over the last two years, and the three year old shrimp show a reduction since 2003 (Table 7.5 and Figure 7.6). As the cod stock is still on a high level, the natural mortality is believed to be high.

The strong 1999 and 2000 year classes did not contribute to the assumed increase in shrimp biomass in 2004. These originally strong year classes seem to have been a target for predators and the shrimp fishery as young shrimp (Figure 7.3).

As the recruitment to the fishery in 2004 is lower than in 2003, the stock is expected to remain at a low level in 2005. Preliminary records indicate a catch of approximately $37,000 \mathrm{t}$ for 2004 , and the stock does not seem to allow higher catches in 2005. The WG recommends, that catches should be maintained at the low current level, until an increase in biomass is detected.

### 7.7.3 Recommendations on further work

- It is highly recommended that the Russian shrimp survey time series is re-established. The lack of Russian survey data is considered a big problem when doing the assessment. It is impossible to evaluate the status of the stock in the Kola Coast and the Goose Bank areas.
- Scientists should evaluate the procedures used in estimating the shrimp consumed by cod;
- Length and sex data from commercial catches should be provided by all nations involved in the fishery;
- Data on discards of small shrimp in the Barents Sea and Svalbard are should be presented in 2005;
- Authorities should enforce the accurate completion of logbook data in Norway, especially the use of single, double and triple trawls;
- Work on developing and evaluating assessment methods should be continued;
- Catch and effort statistics should be delivered to the ICES by all countries involved in the shrimp fishery in the Barents Sea and the Svalbard area. Now only available Norwegian, Russian and Spanish data are available.


## 8 The by-catch in the Pandalus fisheries in the Subarea IV and Division IIIa

### 8.1 Available data

In recent years there has been increasing focus on mixed fisheries or fisheries, where species from stocks subject to recovery plans or under special surveillance. The fisheries for Pandalus in the North Sea area cannot be classified as mixed fisheries as for instance some of the fisheries for Nephrops. The current by-catch regulations in force for the gears used in the fisheries for Pandalus restrict the amounts of by-catch, but nevertheless are several valuable fish species, e.g. cod, anglerfish, taken and landed as by-catch. Since the Pandalus fisheries are classified as 'small mesh fisheries' for 'human consumption (h.c.) species' there has for a long time been concern on the by-catches in these fisheries, and the Pandalus WG has since the 1980s regularly compiled and presented relevant information on by-catch in the WG reports.

Tables 8.1 A - G give for the recent 10 years period the available Danish, Norwegian and Swedish data on bycatch of the main species in the Pandalus fisheries landed for h.c. In the some years quantities of Norway pout and Blue whiting have been specified. For all 3 countries the data are from log book records and are only recording landings, i.e. not the discarded by-catch. Both the Danish and Swedish log book records cover nearly all the recorded Pandalus landings. The Norwegian records for 2002 and 2003 also cover the total landings, while those prior to 2002 only cover approximately $33 \%$ of total landings.

These tables also give cod as well as total h.c. by-catch as the percentage of Pandalus landings. It is believed that these are better estimators than $\%$ of total catch, since log-book recordings probably not always are consistent in recordings of e.g. Norway pout and/or Blue whiting. In Skagerrak the percentages of landed total h.c. by-catch are similar for all 3 countries (excluding trawls with selective grids). Considering cod only, it is noted that the percentage is highest in the Danish fisheries. However, for the Norwegian log-book records it is likely that the rather low percentages of recorded cod is because some of the cod by-catch has not been specified as cod, but merely as unspecified h.c. by-catch. Note that for the Norwegian data the category 'other market fish' is very high compared to this category in the Danish and Swedish data. Note that the Danish by-catches from the Norwegian Deep are higher than the Norwegian. A minor fraction of the Swedish Pandalus fishery is conducted with trawls equipped with a selective grid, and judging from the logbook records of landings by this gear type, it seems to be very efficient in reducing by-catch, see Table 8.1 C and Sect. 8.3.

It cannot be ruled out, that some times in some areas by-catch of valuable species, for instance angler fish, cod and witch flounder is considered a positive contribution to the total landings from a fishing trip for Pandalus.

The current 'at-sea-sampling' programme has provided sporadic samples of discarded by-catch in the Danish and Swedish Pandalus fisheries. However, these data are presently considered to scanty to base any assessments of the amount of e.g. discarded cod on.

### 8.2 The magnitude of cod landings from the Pandalus fisheries.

The historic data given in Tables 8.1 A - G indicate minor fluctuations without any trends in the amount of cod as by-catch. They do not seem to follow the trend in the development of the cod stock in the North Sea and Skagerrak. However, the relative high by-catch figures of Saithe in recent years in contrast to low values for the first half of the 1990s (Denmark and Sweden) could reflect the increase in size of this stock.

These historic cod by-catch figures indicate for instance that in recent years the total of amount of cod landed by the Pandalus fisheries in the North Sea and IIIa by Denmark, Norway and Sweden has fluctuated around 300 t . Since the U.K. shrimp fishery on Fladen Ground has been small in recent years, the overall picture would not change by adding this component. The overall conclusion on the total annual landed by-catch of cod in the Pandalus fisheries in
these areas is that it contributes less than $1 \%$ of total annual landings of cod in the North Sea and Skagerrak. This amount could probably be reduced further, if the shrimp-trawls were equipped with selective grids, as described below.

### 8.3 Improved species selection in shrimp trawls equipped with selective grids.

The current legal minimum mesh size of 35 mm (stretched mesh) in shrimp (Pandalus sp.) trawls implies the catch of also other unwanted undersized fish species and a resulting increase in mortality due to discards. Experiments with species selective grids installed in the trawl started in Norway 1988, and the Nordmøre grid with 20 mm bar space is now mandatory in Norwegian Pandalus trawls in the Norwegian zone. Recent experiments on shrimp fishing grounds in the Norwegian Deeps have shown that the by-catch of cod, haddock, saithe and whiting is low when targeting shrimp at depths deeper than 240-250 meters, which are the common fishing depths in this area. Particularly juveniles of such species are absent in shrimp trawl catches in this fishing area. (Valdemarsen and Misund, 2003). Similar species selective shrimp trawls have been tested in the North Sea and the Skagerrak in an EU Study project by Denmark and Sweden (Madsen et al. 1998). The Swedish experimental fishing was performed both inshore and offshore with identical rigging as in the Norwegian legislation. The results shows that the total proportion of fish in the inshore catch was reduced by $85 \%$ when the Nordmøre grid was used and the remaining fish by-catch consisted almost solely of Norway pout. No significant loss of shrimp could be seen, but average catch of shrimp per trawling hour decreased by about $7 \%$ when using the grid. Even in the offshore fishery the by far largest by-catch was Norway pout, which also is the most difficult species to sort out because of its small size. All other fish species were sorted to $97 \%$, and commercial fish species to $99 \%$ efficacy.

The conclusions from these studies are that an introduction of the equivalent grids in the shrimp trawl fishery will drastically reduce the by-catch of fish in general, and commercial fish species in particular and according to published results, a comparable selection efficacy is unlikely to be achieved using techniques that depend solely on mesh selection.

Detailed description of Nordic experiments with grids in shrimp trawls is found in (Anon., 1996) and an extensive reference list is presented in (ICES, 1998).

## 9 The by-catch in the Pandalus fisheries in the Barents Sea

Young Northeast Arctic cod, haddock, redfish and Greenland halibut are caught as by-catch in shrimp fisheries in the Barents Sea. The figures for the cod by-catch are based on commercial Norwegian commercial shrimp catch statistics, logbook data, surveys and surveillance data from 1983-2002 (Aijad et al. 2004).

Especially one and two year old cod are caught in the shrimp fishery due to overlapping distribution of shrimp and cod in the central area of the Barents Sea and around Svalbard. Cod by-catch in the shrimp fishery has been regulated by area closures since 1983 (Aschan, 1999; Aschan, 2000). Furthermore, a by-catch regulation of max. 3 juveniles of cod and haddock pr 10 kg of shrimp was introduced in 1983. However, with the introduction of the sorting grid in 1995 the number of cod and haddock allowed as by-catch increased to 10 . The results show that the numbers of cod taken as by-catch varied between quarter and between years. Northeast Arctic cod by-catch rates in stay around $1 \%$ of the annual shrimp landing recent years (Table 9.1). There is a declining trend in cod by-catch from 1997-2002. This by-catch rate is not thought to have an essential impact on the recruitment of Northeast Arctic cod (Aijad et al. 2004). The weight and number of individuals of other by-catch species does not exceed the estimates for cod. However strong year classes of haddock may reach the same values as cod.

Ajiad A., Aglen, A. and Nedreaas K. 2004. Cod by-catch motality fom the Barents Sea shrimp fishery 1983-2002. ICES Arctic Fisheries Working Group, WD 24.

Albert, O.T. 1993 Distribution, population structure and diet of silvery pout (Gadiculus argentus thori J. Schmidt), poor cod (Trisopterus minutes minutes (L.)), four-bearded rockling (Rihnonemus cimbrius(L.)), and Vahl's eelpout (Lycodes vahlii gracilis Reinhardt) in the Norwegian Deep. Sarsia 78:141-154.

Albert, O.T. 1994. Ecology of haddock (Melanogrammus aeglefinus L.) in the Norwegian Deep. ICES J. mar. Sci., 51:31-44

Albert, O.T. 1994. Biology and ecology of Norway pout (Trisopterus esmarki Nilson 1855) in the Norwegian Deep. ICES J. mar. Sci., 51:45-61

Anon. 1996. Seleksjon i reketrål. Red. J. W. Valdemarsen. TemaNord 1996:520.

Anon.,2003. ICES, Report Arctic Fisheries Working Group, sect. 9

Aschan, M. 1999. Bioeconomic analyses of by-catch of juvenile fish in the shrimp fisheries. Fiskeriforskning Report 24/1999.

Aschan, M. 2000. Working document on by-catch in the shrimp fishery in the Barents Sea. Fiskeriforskning Preliminary Report, Tromsø, March 2000.

Aschan, M. 2001. Spatial variability in length frequency distribution and growth of shrimp (Pandalus borealis Krøyer 1838) in the Barents Sea. J. Northw. Atl. Fish. Sci. Vol.27: 77-89.

Aschan, M., Berenboim, B., Mukhin, S. and Sunnanå, K. 1993. Results of Norwegian and Russian investigations of shrimp (Pandalus borealis) in the Barents Sea and Svalbard area in 1992. ICES CM 1993/K:9.

Aschan, M., Berenboim, B. and Mukhin, S. 1994. Results of Norwegian and Russian investigations of shrimp (Pandalus borealis) in the Barents Sea and the Svalbard area 1993, compared with earlier studies. ICES CM 1994/K:37.

Aschan, M., Berenboim, B. and Mukhin, S. 1995. Results of Norwegian and Russian investigations of shrimp (Pandalus borealis) in the Barents Sea and the Svalbard area 1994. ICES CM 1995/K:11.

Aschan, M., Berenboim, B. and Mukhin, S. 1996. Results of Norwegian and Russian investigations of Shrimp (Pandalus borealis) in the Barents Sea and in the Svalbard area 1995. ICES CM 1996/K:6.

Aschan, M. and Sunnanå, K. 1997. Evaluation of the Norwegian Shrimp Surveys conducted in the Barents Sea and the Svalbard area 1980-1997. ICES CM 1997/Y:07.

Bakanev S. and Berenboim B. 2004. On the minimal allowable size of the Barents Sea northern shrimp. ICES WGPAND, WD 01: 11 p .

Berenboim, Lysy, Serebrov. 1980 On distribution, stock state and regulation measures of shrimp (Pandalus borealis) fishery in the Barents Sea. ICES C.M.1980/K:15, 18 pp.

Berenboim, B.I., Lysy, A. Yu. and Salmov, V.Z. 1986. Soviet investigations on shrimp (Pandalus borealis) in the Barents Sea and Spitsbergen area in May 1985. ICES CM 1986/K:11.

Berenboim, B.I., Mukhin, S.G. and Sheveleva, G.K. 1989. Soviet investigations of shrimp Pandalus borealis in the Barents Sea and off the Spitsbergen in 1988. ICES CM 1989/K:14.

Berenboim, B.I., Mukhin, S.G. and Sheveleva, G.K. 1990. Soviet investigations of shrimp Pandalus borealis in the Barents Sea and off the Spitsbergen in 1989. ICES CM 1990/K: 4.

Berenboim, B., Mukhin, S. and Sunnanå, K. 1992. Results from Norwegian and Soviet investigations of shrimp Pandalus borealis in the Barents Sea. ICES CM 1992/K:14.

Berenboim, B. and Korzhev, V. 1997. On possibility of using Stefansson's production model to assess the northern shrimp (Pandalus borealis) stock in the Barents Sea. ICES CM 1997/Y.

Berenboim, B.I., Dolgov, A.V., Korzhev, V.A. and Yaragina N.A. 2001. The impact of cod on the dynamics of Barents Sea shrimp (Pandalus borealis) as determined by multispecies models. J. Northw. Atl. Fish. Sci. 27: 1-7.

Bergstad, O.A. 1991. Distribution and trophic ecology of some gadoid fish of the Norwegian Deep. 1. Accounts of individual species. Sarsia 75:269-313

Bhattacharya C. G., 1967. A simple method of resolution of a distribution into Gaussian components. Biometrics, 23:115-135

Bogstad, B. and Mehl, S. (MS) 1997. Interactions Between Cod (Gadus morhua) and Its Prey Species in the Barents Sea. Forage Fishes in Marine Ecosystems. Proceedings of the International Symposium on the Role of Forage Fishes in Marine Ecosystems. Alaska Sea Grant College Program Report No. 97-01: 591-615. University of Alaska Fairbanks.

Bulgakova, T., D.Vasilyev, V.Korzhev, V.Tretyak 1995. The results of multispecies analysis for the Barents Sea fishery community (cod, capelin, shrimp and herring). ICES C.M. 1995/D:14.

Caddy, J.F. 1999. Deciding on precautionary management measures for a stock based on a suite of limit reference points (LRPs) as a basis for a multi-LRP harvest law. NAFO Sci. Coun. Studies, 32: 55-68.

Cadrin, S.X. 2000 Evaluating two assessment methods for Gulf of Maine northern shrimp based on simulations.
J.Northwest. Atl. Fish. Sci., vol. 27:119-132

Dolgov A.V. 1997. Distribution, abundance, biomass and feeding of thorny skate, Raja radiata, in the Barents Sea. ICES CM 1997/GG: 04.

Dolgova N.V and Dolgov A.V. 1997. Stock status and predation of long rough dab (Hippoglossoides platessoides) in the Barents and Norwegian Seas. (In: International Symposium on the Role of Forage Fishes in Marine Ecosystems, Anchorage, Alaska 13-16 November 1996.

Drengstig, A., Fevolden S.E, Galand P.E, and Aschan, M.M. 2000. Population structuring of the deep sea shrimp (Pandalus borealis) in the NE Atlantic based on allozymic differentiation. Aquat. Living Resour. 13:1-9.

Fournier D.A., Sibert J.R., Majkowski J. and Hampton J., 1990. MULTIFAN a likelihood-based method for estimating growth parameters and age composition from multiple length frequency data sets illustrated using data for Southern Bluefin Tuna (Thunnus maccoyii). Can. J. Fish. Aquat. Sci., 47, 1990

Hansen, H.Ø., and Aschan, M. 2001. Growth performance, size and age at maturity of shrimp Pandalus borealis, in the Svalbard area related to environment. J. Northw. Atl. Fish. Sci.,27:83-92.

Harbitz, A., Aschan, M. and Sunnanå, K. 1998. Optimum stratified sampling design for biomass estimates in large area trawl surveys - exemplified by shrimp surveys in the Barents Sea. Fisheries research 37:107-113.

Hylen, A., Tveranger, B. and Øynes, P. 1984. Norwegian investigations on the deep Sea shrimp (Pandalus borealis) in the Barents Sea in April - May 1984 and in the Spitsbergen area in July - August 1984. ICES CM 1984/K:21.

Hylen, A. and Øynes, P. 1986. Results of stratified trawl surveys for shrimp (Pandalus borealis) in the Barents Sea and in the Svalbard region in 1986. ICES CM 1986/K:34.

Hylen, A., Jacobsen, J. A. and Øynes, P. 1987. Results of stratified trawl surveys for shrimp (Pandalus borealis) in the Barents Sea and the Svalbard region in 1987. ICES CM 1987/K:39.

Hylen, A. and Øynes, P. 1988. Results of stratified trawl surveys for shrimp (Pandalus borealis) in the Barents Sea and the Svalbard region in 1988. ICES CM 1988/K:18.

Hylen, A., Sunnanå, K. and Øynes, P. 1989. Results of stratified trawl surveys for shrimp (Pandalus borealis) in the Barents Sea and the Svalbard region in 1989. ICES CM 1989/K:26.

Hylen, A. and Ågotnes, P. 1990. Results of stratified shrimp trawl surveys for shrimp (Pandalus borealis) in the Barents Sea and the Svalbard region in 1990. Survey Report.

ICES (1990). Report on the Working group on the assessment of Pandalus Stocks ICES C.M.1990/Assess:9.
ICES (1992). Report on the Working group on Nephrops and Pandalus Stocks. ICES C.M.1992/Assess:8.

ICES 1998. Report of the Study Group on grid (grate) sorting systems in trawls, beamtrawls and seine nets. ICES C.M. 1998/B:2

ICES (2000). Report on the Pandalus Assessment Working group. ICES C.M.2000/ACFM:2.
ICES (2001). Report on the Pandalus Assessment Working group, 2001. ICES C.M.2002/ACFM:4
ICES (2003). Report on the Pandalus Assessment Working group, 2003. ICES C.M.2004/ACFM:5
ICES (2002b) Report of the Study Group on the Further Development of the Precautionary Approach to Fishery
Management. ICES CM 2002/ ACFM:10

Kartavtsev, V.P., Berenboim, B. and Zugurovsky, K.I. 1991. Population genetic differentiation on the pink shrimp Pandalus borealis Krøyer 1838, from the Barents and Bering Seas. J. Shell. Res. 10 (2): 333-339.

Koeller, P., Savard, L., Parsons, D.G. and Fu, C. 2001. A precautionary approach to assessment and management of shrimp stocks in the Northwest Atlantic. J. Northw. Atl. Fish. Sci. Vol.27: 235-246.

Kristjansson, J. 1994 Geometric measurements on the Campelen 1800 sampling trawl at resource surveys in the Barent Sea and in the Svalbard area 1992-1993. Experiences with blocked door width. Fiskeriforskning report. 18 pp. [In Norwegian].

Macdonald, P.D.M., and T.J. Pitcher. 1979. Age groups from size-frequency data: a versatile and efficient method of analysing distribution mixtures. J. Fish. Res. Board Can., 36: 987-1001.

Madsen, N., P-O. Larsson and M. Ulmestrand, 1998. Development end testing of grids for the Skagerrak and North Sea shrimp fishery. Interim Report. EU Study contract No. 96/004

Martinez, I., Skjeldal, T.O., Dreyer, B. and Aljanabi, S.M. 1997. Genetic structuring of Pandalus borealis in the NEAtlantic. II. RAPD analysis. ICES CM 1997/T:24.

Mukhin, S.G. and Sheveleva, G.K. 1991. Soviet investigations on shrimp in the Barents Sea and off the Spitsbergen in 1990. ICES CM 1991/K:14.

Patterson, K., 1992: Fisheries for small pelagic species: an empirical approach to management targets. Reviews in Fish Biology and Fisheries, 2, 321-338 (1992).

Pedersen 2003, O. P., Aschan, M., Te, K., Slagstad, D. and Rasmussen, T. The advection and population dynamics of Pandalus borealis investigated by Lagrangian particle tracking model. Fisheries Research, 65:173-190.

Poulsen E.M., 1970. On deep-sea prawn in the North Sea-Skagerrak. Medd. Dan. Fish. havunders. 7(1):22
dos Santos, J and Jobling, M. 1995. Test of a food consumption model for the Atlantic cod. ICES J. mar. Sci., 52:209219.

Rasmussen, T., Thollesson, M. and Nilssen, E.M. 1993. Preliminary investigations on the population genetic differentiation of the deep water prawn, Pandalus borealis Krøyer 1838, from Northern Norway and the Barents Sea. ICES CM 1993/K:11.

Skjæraasen, J.E. 1998 Skater (Pisces: Rajidae) i Norskerenna og langs norsk kontinentalskråning: utbredelse, ernæring og biologi. Thesis 89 pp

Sparre P. 1984. A computer program for estimation of food suitability coefficients from stomach content data and multispecies VPA. ICES C.M. 1984/G:25.

Stefánsson, G., U. Skúladóttir and G. Pétursson, 1994. The use of a stock production type model in evaluating the offshore Pandalus borealis stock of North Icelandic waters, including the predation of Northern Shrimp by Cod. - ICES C.M. 1994/K:25, 13 pp .

Tavares, A. M. and Øynes, P. 1980. Results of a stratified bottom trawl survey for shrimp (Pandalus borealis) in the Barents Sea and the Spitsbergen area in May-June 1980. ICES CM 1980/K:22.

Teigsmark, G. and Øynes, P. 1981. Results of a stratified bottom trawl survey for shrimp (Pandalus borealis) in the Barents Sea in May - June 1981. ICES CM 1981/K:21.

Teigsmark, G. and Øynes, P. 1982. Norwegian investigations on the deep Sea shrimp (Pandalus borealis) in the Barents Sea in 1982. ICES CM 1982/K:12.

Teigsmark, G. and Øynes, P. 1983a. Results of a stratified bottom trawl survey for shrimp (Pandalus borealis) in the Spitsbergen area in July 1982. ICES CM 1983/K:17.

Teigsmark, G. and Øynes, P. 1983b. Norwegian investigations on the deep Sea shrimp (Pandalus borealis) in the Barents Sea in April - May 1983 and in the Spitsbergen area in July 1983. ICES CM 1983/K:46.

Tveranger, P. and Øynes, P. 1985. Results of stratified trawl surveys for shrimp (Pandalus borealis) in the Barents Sea in May and in the Svalbard region in July - August 1985. ICES CM 1985/K:50.

Torres,P., E.Roderiguez-Marin and I. Loureiro 2000. Preliminary results from feeding analysis for the most abundant demersal fishes in Flemish Cap during summer (1993-2000). NAFO SCR Doc 00/60 9pp.

Valdemarsen, J.W. and Misund, R., 2003. Forsøk med 19 og 22 mm spileavstand i sorteringsrist i fisket etter rognreke i Nordsjøen våren 2002. Report. Institute of Marine Research/Directorate of Fisheries, Bergen, Norway.

Ulrich, C., and Andersen B. S. 2004. Dynamics of Fisheries, and the flexibility of vessel activity in denmark between 1989 and 2001. - Ices Journal of Marine Science, 61: 308-322.


Tabel 4.2 Pandalus borealis landings from divisions IIIa (Skagerrak) and IVa (eastern part). as estimated by the Working Group

| Year | Denmark | Norway | Sweden | Total | stimated discards | TAC | Catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1970 | 1102 | 1729 | 2742 | 5573 |  |  |  |
| 1971 | 1190 | 2486 | 2906 | 6582 |  |  |  |
| 1972 | 1017 | 2477 | 2524 | 6018 |  |  |  |
| 1973 | 755 | 2333 | 2130 | 5218 |  |  |  |
| 1974 | 530 | 1809 | 2003 | 4342 |  |  |  |
| 1975 | 817 | 2339 | 2003 | 5159 |  |  |  |
| 1976 | 1204 | 3348 | 2529 | 7081 |  |  |  |
| 1977 | 1120 | 3004 | 2019 | 6143 |  |  |  |
| 1978 | 1459 | 2440 | 1609 | 5508 |  |  |  |
| 1979 | 1062 | 3040 | 1787 | 5889 |  |  |  |
| 1980 | 1678 | 4562 | 2159 | 8399 |  |  |  |
| 1981 | 2593 | 5183 | 2241 | 10017 |  |  |  |
| 1982 | 3766 | 5042 | 1450 | 10258 |  |  |  |
| 1983 | 1567 | 5361 | 1136 | 8064 |  |  |  |
| 1984 | 1800 | 4783 | 1022 | 7605 | 200 |  | 7805 |
| 1985 | 4498 | 6646 | 1571 | 12715 | 558 |  | 13273 |
| 1986 | 4866 | 6490 | 1463 | 12819 | 414 |  | 13233 |
| 1987 | 4488 | 8343 | 1322 | 14153 | 723 |  | 14876 |
| 1988 | 3240 | 7661 | 1278 | 12179 | 750 |  | 12929 |
| 1989 | 3242 | 6411 | 1433 | 11086 | 1107 |  | 12193 |
| 1990 | 2479 | 6108 | 1608 | 10195 | 1226 |  | 11421 |
| 1991 | 3583 | 6119 | 1908 | 11610 | 497 |  | 12107 |
| 1992 | 3725 | 7136 | 2154 | 13015 | 541 | 15000 | 13556 |
| 1993 | 2915 | 7371 | 2300 | 12586 | 889 | 15000 | 13475 |
| 1994 | 2134 | 6813 | 2601 | 11548 | 214 | 18000 | 11761 |
| 1995 | 2460 | 8095 | 2882 | 13437 | 275 | 16000 | 13713 |
| 1996 | 3868 | 7878 | 2371 | 14117 | 318 | 15000 | 14436 |
| 1997 | 3909 | 8565 | 2597 | 15071 | 1039 | 15000 | 16110 |
| 1998 | 3330 | 9606 | 2469 | 15406 | 348 | 18800 | 15753 |
| 1999 | 2072 | 6739 | 2445 | 11256 | 639 | 18800 | 11895 |
| 2000 | 2371 | 6118 | 2225 | 10714 | 687 | 13000 | 11401 |
| 2001 | 1953 | 6895 | 2108 | 10956 | 701 | 14500 | 11657 |
| 2002 | 2466 | 7321 | 2301 | 12088 | 254 | 14500 | 12342 |
| 2003 | 3244 | 7715 | 2389 | 13348 | *) |  | 13348 *) |

[^1]| Table 4.3 National LPUE and total effort as estimated by the Working Group, Pandalus division IIIa and IVa east |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Denmark | Total | Norway | Total | Sweden | Total | average index |
|  | LPUE | effort | LPUE | effort | LPUE | effort |  |
|  | kg/day | days | kg/hr | Khrs | kg/hr | Khrs |  |
| 1984 | 452 | 3869 |  |  | 25 | 40 |  |
| 1985 | 743 | 6053 |  |  | 32 | 49 |  |
| 1986 | 556 | 8700 | 36 | 179 | 30 | 49 | 1.00 |
| 1987 | 499 | 9212 | 36 | 230 | 23 | 57 | 1.17 |
| 1988 | 432 | 7104 | 31 | 251 | 22 | 57 | 1.13 |
| 1989 | 441 | 7143 | 23 | 273 | 23 | 63 | 1.21 |
| 1990 | 591 | 4195 | 26 | 232 | 26 | 58 | 0.99 |
| 1991 | 645 | 5555 | 30 | 206 | 31 | 61 | 1.01 |
| 1992 | 641 | 5811 | 35 | 204 | 27 | 80 | 1.15 |
| 1993 | 571 | 5068 | 31 | 243 | 25 | 91 | 1.27 |
| 1994 | 677 | 3146 | 31 | 218 | 33 | 82 | 1.08 |
| 1995 | 801 | 3072 | 35 | 255 | 39 | 76 | 1.11 |
| 1996 | 860 | 4466 | 37 | 214 | 32 | 74 | 1.07 |
| 1997 | 1034 | 3770 | 42 | 212 | 33 | 78 | 1.07 |
| 1998 | 1023 | 3256 | 44 | 219 | 34 | 73 | 1.03 |
| 1999 | 833 | 2501 | 32 | 219 | 34 | 72 | 0.99 |
| 2000 | 870 | 2713 | 31 | 195 | 30 | 75 | 0.98 |
| 2001 | 840 | 2314 | 32 | 217 | 29 | 74 | 1.00 |
| 2002 | 1069 | 2306 | 39 | 186 | 35 | 65 | 0.88 |
| 2003 | 1073 | 3013 | 47 | 166 | 33 | 72 | 0.91 |


| Table 4.4 Total international LPUE and effort as estimated by the Working Group , |  |  |
| :--- | :---: | :---: |
| Year | RPUE | effort <br> Khrs |
| 1984 | 22.0 | 345 |
| 1985 | 32.8 | 388 |
| 1986 | 30.3 | 424 |
| 1987 | 28.6 | 494 |
| 1988 | 26.0 | 468 |
| 1989 | 22.3 | 497 |
| 1990 | 26.5 | 384 |
| 1991 | 29.6 | 392 |
| 1992 | 31.4 | 415 |
| 1993 | 28.1 | 448 |
| 1994 | 31.1 | 371 |
| 1995 | 33.6 | 400 |
| 1996 | 36.3 | 388 |
| 1997 | 40.2 | 375 |
| 1998 | 42.2 | 365 |
| 1999 | 32.4 | 347 |
| 2000 | 32.4 | 331 |
| 2001 | 31.9 | 343 |
| 2002 | 39.9 | 303 |
| 2003 | 43.7 | 306 |



| Table 4.6 Catch in numbers at age. Pandalus division Illa and IVa east. |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Numbers*10**-6 |  |  |  |  |  |  |  |  |  |  |
| YEAR | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 | 17.7 | 7.4 | 2.7 | 14.1 | 31.3 | 0.0 | 3.9 | 25.5 | 27.2 | 0.7 |
| 1 | 1200.8 | 1146.4 | 1260.5 | 1086.6 | 2083.6 | 2250.1 | 1231.8 | 1071.4 | 1889.6 | 671.9 |
| 2 | 1305.4 | 1029.7 | 1205.6 | 923.9 | 385.5 | 910.8 | 1035.8 | 1289.2 | 803.8 | 1380.4 |
| 3 | 187.9 | 482.7 | 390.2 | 300.2 | 173.8 | 121.1 | 326.7 | 569.1 | 262.7 | 143.0 |
| +gp | 52.3 | 25.1 | 203.2 | 146.7 | 13.6 | 31.3 | 25.6 | 57.5 | 15.5 | 30.5 |
| TOTALNUM | 2764.1 | 2691.3 | 3062.1 | 2471.5 | 2687.9 | 3313.3 | 2623.8 | 3012.7 | 2998.7 | 2226.4 |
| TONSLAND | 13273 | 13233 | 14876 | 12929 | 12193 | 11421 | 12107 | 13556 | 13475 | 11761 |
| SOPCOF\% | 89 | 97 | 105 | 102 | 106 | 88 | 97 | 88 | 93 | 0 |
| YEAR | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 | 2.7 | 61.1 | 19.7 | 12.7 | 4.6 | 88.1 | 0.0 | 3.9 | 2.4 |  |
| 1 | 646.0 | 1211.6 | 2175.6 | 903.4 | 1436.1 | 1270.7 | 1308.0 | 922.3 | 668.7 |  |
| 2 | 970.5 | 991.4 | 1181.9 | 1597.9 | 720.1 | 836.3 | 826.2 | 858.4 | 1466.5 |  |
| 3 | 851.5 | 454.6 | 295.6 | 468.1 | 318.3 | 199.3 | 382.5 | 581.8 | 283.8 |  |
| +gp | 42.0 | 69.5 | 29.8 | 48.2 | 43.3 | 39.2 | 80.8 | 101.8 | 0.0 |  |
| TOTALNUM | 2512.5 | 2788.2 | 3702.6 | 3030.2 | 2522.4 | 2433.5 | 2597.5 | 2468.3 | 2421.4 |  |
| TONSLAND | 13713 | 14436 | 16110 | 15753 | 11895 | 11401 | 11657 | 12339 | 13338 |  |
| SOPCOF\% | 87 | 88 | 94 | 96 | 95 | 95 | 90 | 88 | 99 |  |


| Table 4.7 Mean weight at age in catches. Pandalus division IIla and IVa east. |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Catchweights at age (kg) |  |  |  |  |  |  |  |  |  |  |
| YEAR | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 | 0.0009 | 0.0012 | 0.0009 | 0.0009 | 0.0011 | 0.0009 | 0.0015 | 0.0010 | 0.0009 | 0.0009 |
| 1 | 0.0032 | 0.0032 | 0.0024 | 0.0030 | 0.0034 | 0.0030 | 0.0033 | 0.0035 | 0.0035 | 0.0034 |
| 2 | 0.0064 | 0.0054 | 0.0048 | 0.0054 | 0.0065 | 0.0053 | 0.0053 | 0.0052 | 0.0067 | 0.0060 |
| 3 | 0.0104 | 0.0083 | 0.0077 | 0.0090 | 0.0099 | 0.0083 | 0.0079 | 0.0078 | 0.0088 | 0.0093 |
| +gp | 0.0134 | 0.0140 | 0.0114 | 0.0117 | 0.0133 | 0.0106 | 0.0122 | 0.0095 | 0.0109 | 0.0117 |
| YEAR | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 |  |
| AGE |  |  |  |  |  |  |  |  |  |  |
| 0 | 0.0009 | 0.0007 | 0.0009 | 0.0007 | 0.0007 | 0.0007 | 0.0006 | 0.0008 | 0.0014 |  |
| 1 | 0.0033 | 0.0037 | 0.0031 | 0.0033 | 0.0033 | 0.0032 | 0.0031 | 0.0036 | 0.0035 |  |
| 2 | 0.0057 | 0.0067 | 0.0061 | 0.0055 | 0.0063 | 0.0063 | 0.0056 | 0.0054 | 0.0060 |  |
| 3 | 0.0089 | 0.0094 | 0.0094 | 0.0087 | 0.0088 | 0.0103 | 0.0086 | 0.0083 | 0.0082 |  |
| +gp | 0.0116 | 0.0138 | 0.0119 | 0.0133 | 0.0112 | 0.0139 | 0.0117 | 0.0113 | 0.0121 |  |


| Table 4.8 | Norwegia Numbers | shrimp illions) | $\begin{aligned} & \text { urvey i } \\ & \text { t age } r \end{aligned}$ | $\mathrm{a} \& \mathrm{IV}$ ed to | Catch | Pandalus. area (1-16) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Age |  |  |  |
| Year | 0 | 1 | 2 | 3 | 4 | 5 |
| 1984 | 307 | 2000 | 844 | 348 | 45 | 1 |
| 1985 | 269 | 3574 | 1635 | 323 | 144 | 1 |
| 1986 | 130 | 1046 | 685 | 282 | 20 | 0 |
| 1987 | 91 | 2600 | 1105 | 729 | 93 | 0 |
| 1988 | 220 | 337 | 415 | 294 | 53 | 1 |
| 1989 | 993 | 1951 | 447 | 166 | 7 | 0 |
| 1990 | 455 | 1850 | 919 | 98 | 7 | 0 |
| 1991 | 224 | 2521 | 996 | 211 | 26 | 0 |
| 1992 | 2264 | 1906 | 1107 | 423 | 38 | 0 |
| 1993 | 476 | 3075 | 890 | 332 | 17 | 0 |
| 1994 | 267 | 1862 | 1024 | 413 | 136 | 2 |
| 1995 | 170 | 1384 | 759 | 929 | 37 | 0 |
| 1996 | 915 | 2827 | 1204 | 538 | 42 | 0 |
| 1997 | 225 | 3474 | 1696 | 715 | 313 | 0 |
| 1998 | 331 | 1096 | 1375 | 1027 | 159 | 0 |
| 1999 | 309 | 1938 | 529 | 716 | 45 | 0 |
| 2000 | 838 | 2319 | 900 | 296 | 94 | 5 |
| 2001 | 737 | 2630 | 1260 | 795 | 124 | 0 |
| 2002 | 494 | 3431 | 1116 | 653 | 129 | 1 |
| 2003 | 179 | 1144 | 346 | 755 | 424 | 2 |
| 2004 | n.a. | 5657 | 1971 | 1207 | 390 | 33 |


| Table 4.9 | Shrimp (tonnes) raised to areas, assuming catch efficiency 1.0 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | $\begin{aligned} & \hline \text { Input to } \\ & \text { SPP *) } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stratum nr. | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | Total |  |
| Area in $N M^{\wedge}$ | 2794 | 1819 | 342 | 1824 | 1559 | 1392 | 860 | 206 | 431 | 299 | 400 | 295 | 561 | 1685 | 957 | 1136 | 16560 |  |
| Year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1984 | 0 | 2441 |  | 2144 | 4048 | 3093 | 1313 |  | 336 | 346 | 316 | 556 | 605 | 1253 | 1305 | 1535 | 19291 | 19291 |
| 1985 | 0 | 4768 |  | 1162 | 3288 | 2607 | 2016 | 0 | 815 | 475 | 1900 | 794 | 840 | 4921 | 2664 | 4066 | 30316 | 30316 |
| 1986 | 0 | 2183 |  | 920 | 933 | 1940 | 663 |  | 389 | 177 | 857 | 540 | 618 | 1521 | 2073 | 733 | 13547 | 13547 |
| 1987 | 88 | 3765 |  | 2482 | 4103 | 3294 | 1237 | 0 | 1370 | 254 | 1470 | 584 | 419 | 2168 | 1350 | 964 | 23548 | 23460 |
| 1988 | 0 | 1126 |  | 720 | 373 | 1079 | 682 | 0 | 294 | 96 | 472 | 391 | 282 | 814 | 777 | 343 | 7449 | 7449 |
| 1989 |  | 932 |  | 2347 | 898 | 1722 | 1159 | 0 | 560 | 263 | 579 | 556 | 498 | 1375 | 1443 | 918 | 13248 | 13248 |
| 1990 | 0 | 705 | 187 | 3245 | 1067 | 2373 | 471 | 0 | 647 | 171 | 1044 | 559 | 564 | 2088 | 1895 | 907 | 15920 | 15733 |
| 1991 | 0 | 1903 | 1008 | 2612 | 189 | 2851 | 1053 | 152 | 725 | 189 | 740 | 526 | 716 | 2163 | 2683 | 1312 | 18821 | 17661 |
| 1992 | 0 | 615 | 717 | 585 | 136 | 5743 | 2299 | 0 | 568 | 527 | 2091 | 951 | 669 | 3567 | 2550 | 1211 | 22229 | 21512 |
| 1993 | 0 | 1481 | 401 | 4063 | 1487 | 1437 | 688 |  | 621 | 281 | 2596 | 758 | 728 | 2735 | 3823 | 1237 | 22336 | 21935 |
| 1994 | 0 | 1391 | 626 | 2321 | 345 | 2439 | 1992 |  | 461 | 255 | 1627 | 468 | 844 | 3004 | 2284 | 1320 | 19377 | 18751 |
| 1995 | 0 | 2794 |  | 1420 | 202 | 4042 | 953 |  | 818 | 236 | 1836 | 513 | 665 | 2950 | 2076 | 1714 | 20220 | 20220 |
| 1996 | 0 | 4901 |  | 1367 | 133 | 3576 | 1108 |  | 533 | 441 | 3590 | 616 | 921 | 4277 | 2456 | 1286 | 25205 | 25205 |
| 1997 | 0 | 7882 |  | 1995 | 416 | 3393 | 2406 |  | 764 | 349 | 1969 | 1530 | 1487 | 3199 | 3584 | 3169 | 32143 | 32143 |
| 1998 |  | 5069 |  | 3357 | 586 | 2223 | 1049 |  | 682 | 401 | 1105 | 451 | 529 | 3186 | 2439 | 1378 | 22455 | 22455 |
| 1999 | 0 | 5180 |  | 5360 | 3158 | 3254 | 1051 |  | 235 | 243 | 475 | 266 | 311 | 4560 | 2228 | 1596 | 27917 | 27917 |
| 2000 |  | 3436 |  | 2664 | 1121 | 2181 | 695 |  | 343 | 158 | 939 | 380 | 286 | 4159 | 2495 | 1497 | 20354 | 20354 |
| 2001 |  | 5180 | 0 | 5360 | 3158 | 3254 | 1051 |  | 307 | 245 | 512 | 266 | 311 | 4560 | 2228 | 1596 | 28028 | 28028 |
| 2002 |  | 3922.1 |  | 3104 | 459 | 3749 | 1847 |  | 1153 | 364 | 1403 | 496 | 411 | 5425 | 4470 | 3329 | 30133 | 30133 |
| 2003**) |  |  |  | 1410 | 750 | 2770 | 840 | 300 | 1240 | 430 | 480 | 770 | 960 | 2210 | 1950 | 850 | 15070 | 14660 |
| 2004**) |  | 3590 |  | 2830 |  | 3540 | 1530 |  | 690 | 400 | 120 | 1390 | 1230 | 11060 | 4650 | 2890 | 33920 | 33920 |
| *) Areas 1,3 and 8 were exluded when biomass indices for the SPP was calculated. ! The value in the greyed cells are estimated from row- and column means ! |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


*) Based on weights in the catches (WECA)
$\left.{ }^{*}\right)$ The results from the surveys conducted in 2003 and 2004 are not directly comparable with the previous surveys, see Sect. 4.4.1

| Table 4.12 Input data for the SPP model (2003 assessment) |  |  |  |
| :---: | :---: | :---: | :---: |
| Year | Catch tonnes | Recr. Index tonnes | Predator Index kg /NM |
| 1985 | 13273 | 7573 | 40.9 |
| 1986 | 13233 | 13549 | 51.7 |
| 1987 | 14876 | 3327 | 43.9 |
| 1988 | 12929 | 6272 | 42.4 |
| 1989 | 12193 | 1011 | 28.6 |
| 1990 | 11421 | 6586 | 41.2 |
| 1991 | 12107 | 7453 | 35.5 |
| 1992 | 13556 | 9289 | 34.4 |
| 1993 | 13475 | 8837 | 44.2 |
| 1994 | 11761 | 12358 | 45.6 |
| 1995 | 13713 | 8038 | 47.0 |
| 1996 | 14436 | 5903 | 58.8 |
| 1997 | 16110 | 11993 | 41.5 |
| 1998 | 15753 | 13453 | 48.3 |
| 1999 | 11895 | 4667 | 63.1 |
| 2000 | 11401 | 8159 | 50.9 |
| 2001 | 11665 | 10207 | 40.1 |
| 2002 | 12339 | 10870 | 36.0 |
| Mean 85-02 | 13119 | 8308 | 44 |
| Catches for Div. Illa+IVaE . <br> Recruitment index from the Norwegian Trawl Surveys, weight of 1-gr shrimp <br> Predation index as weight of 23 species per nautical mile. |  |  |  |


| Table 4.13 | Biomass indic (2003 assessm Total biomass | es from surveys an ment) | d output from | model |
| :---: | :---: | :---: | :---: | :---: |
| Year | Total biomass tonnes | Obs. biomass index tonnes* $10^{-3}$ | alc. biom index tonnes* $10^{-3}$ | rimps eaten tonnes |
| 1985 | 89996 | 11.4 | 11.1 | 27292 |
| 1986 | 91493 | 16.5 | 11.3 | 34487 |
| 1987 | 119848 | 10.1 | 14.8 | 29275 |
| 1988 | 93226 | 17.1 | 11.5 | 28271 |
| 1989 | 86642 | 6.1 | 10.7 | 19070 |
| 1990 | 60110 | 5.3 | 7.4 | 27527 |
| 1991 | 57958 | 7.6 | 7.2 | 23719 |
| 1992 | 63891 | 8.0 | 7.9 | 22974 |
| 1993 | 79503 | 9.1 | 9.8 | 29481 |
| 1994 | 85933 | 9.0 | 10.6 | 30423 |
| 1995 | 113109 | 10.4 | 14.0 | 31359 |
| 1996 | 112478 | 14.1 | 13.9 | 39252 |
| 1997 | 91080 | 12.0 | 11.2 | 27720 |
| 1998 | 114471 | 18.4 | 14.1 | 32240 |
| 1999 | 141738 | 17.3 | 17.5 | 42101 |
| 2000 | 112643 | 19.3 | 13.9 | 33963 |
| 2001 | 112409 | 8.9 | 13.9 | 26767 |
| 2002 | 130774 | 16.2 | 16.1 | 23999 |
| Mean 85-02 | 97628 | 12 | 12 | 29440 |


| Table 4.14 | Proxies for F and M from the SPP model (2003 assessment) |  |  |
| :---: | :---: | :---: | :---: |
| Year | $\mathrm{F} \approx \mathrm{Y} / \mathrm{B}$ | M $\sim$ Shrimp Eaten/B | "F/M" |
|  | \% | \% | \% |
| 1985 | 14.6 | 30.3 | 48.2 |
| 1986 | 12.5 | 37.7 | 33.2 |
| 1987 | 14.0 | 24.4 | 57.2 |
| 1988 | 14.4 | 30.3 | 47.4 |
| 1989 | 16.6 | 22.0 | 75.5 |
| 1990 | 19.3 | 45.8 | 42.2 |
| 1991 | 19.9 | 40.9 | 48.6 |
| 1992 | 18.9 | 36.0 | 52.6 |
| 1993 | 16.3 | 37.1 | 43.9 |
| 1994 | 11.8 | 35.4 | 33.4 |
| 1995 | 12.2 | 27.7 | 43.8 |
| 1996 | 14.2 | 34.9 | 40.6 |
| 1997 | 15.7 | 30.4 | 51.5 |
| 1998 | 12.3 | 28.2 | 43.7 |
| 1999 | 9.4 | 29.7 | 31.5 |
| 2000 | 10.1 | 30.2 | 33.6 |
| 2001 | 9.6 | 23.8 | 40.3 |
| 2002 | 8.6 | 18.4 | 47.1 |
| Mean 85-02 | 13.9 | 31.3 | 45.2 |


| Table 4.15 | Comparison between 'swept-area' estimates of total biomass and estimates from the SPP model when $B_{0}$ equals 90000 ton |  |  |
| :---: | :---: | :---: | :---: |
|  | Estimated biomass 'available to trawl' (total survey Area) | Estimated total biomass <br> (total survey Area) | $\begin{aligned} & \text { SSP estimates } \\ & \mathrm{B}_{0}=90000 \mathrm{t} \end{aligned}$ |
| 1985 | 19291 | 96456 | 89996 |
| 1986 | 30316 | 151580 | 91493 |
| 1987 | 13547 | 67737 | 119848 |
| 1988 | 23460 | 117298 | 93226 |
| 1989 | 7449 | 37247 | 86642 |
| 1990 | 13248 | 66242 | 60110 |
| 1991 | 15733 | 78665 | 57958 |
| 1992 | 17661 | 88306 | 63891 |
| 1993 | 21512 | 107560 | 79503 |
| 1994 | 21935 | 109673 | 85933 |
| 1995 | 18751 | 93756 | 113109 |
| 1996 | 20220 | 101101 | 112478 |
| 1997 | 25205 | 126025 | 91080 |
| 1998 | 32143 | 160714 | 114471 |
| 1999 | 22455 | 112276 | 141738 |
| 2000 | 27917 | 139587 | 112643 |
| 2001 | 20354 | 101771 | 112409 |
| 2002 | 28028 | 140140 | 130774 |
| 2003 | 30133 | 150666 | 154791 |

Table 5.1 Landings in tonnes of Pandalus borealis from the Fladen Ground (Division IVa) as estimated by the Working Group

| Year | Denmark | Norway | Sweden | UK (Scotland) | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 2204 |  |  | 187 | 2391 |
| 1973 | 157 |  |  | 163 | 320 |
| 1974 | 282 |  |  | 434 | 716 |
| 1975 | 1308 |  |  | 525 | 1833 |
| 1976 | 1552 |  |  | 1937 | 3489 |
| 1977 | 425 | 112 |  | 1692 | 2229 |
| 1978 | 890 | 81 |  | 2027 | 2998 |
| 1979 | 565 | 44 |  | 268 | 877 |
| 1980 | 1122 | 76 |  | 377 | 1575 |
| 1981 | 685 | 1 |  | 347 | 1033 |
| 1982 | 283 |  |  | 352 | 635 |
| 1983 | 5729 | 8 |  | 1827 | 7564 |
| 1984 | 4553 | 13 |  | 25 | 4591 |
| 1985 | 4188 |  |  | 1341 | 5529 |
| 1986 | 3416 |  |  | 301 | 3717 |
| 1987 | 8620 |  |  | 686 | 9306 |
| 1988 | 1662 | 2 |  | 84 | 1748 |
| 1989 | 2495 | 25 |  | 547 | 3067 |
| 1990 | 1681 | 3 | 4 | 365 | 2053 |
| 1991 | 422 | 31 |  | 53 | 506 |
| 1992 | 1448 |  |  | 116 | 1564 |
| 1993 | 1521 | 38 |  | 509 | 2068 |
| 1994 | 1229 | 0 |  | 35 | 1264 |
| 1995 | 4659 | 15 |  | 1298 | 5972 |
| 1996 | 3858 | 32 |  | 1893 | 5783 |
| 1997 | 3022 | 9 |  | 365 | 3396 |
| 1998 | 2900 | 3 |  | 1365 | 4268 |
| 1999 | 1005 | 9 |  | 456 | 1470 |
| 2000 | 1482 |  |  | 378 | 1860 |
| 2001 | 1263 | 18 |  | 397 | 1678 |
| 2002 | 1147 | 9 |  | 70 | 1226 |
| 2003 | 999 | 8 | 1 | 0 | 1008 |
| Note: | 2003 figur | are prelin | inary. |  |  |

Table 5.2 Pandalus borealis, Fladen Ground. Reported LPUE (shrimp trawlers), and estimated total effort.


Table 6.1 Landings (t) of Pandalus borealis from division IVb, the Farn Deeps as estimated by the Working Group

| Year | UK (England) | UK (Scotland) | Denmark | Total |
| :---: | :---: | :---: | :---: | :---: |
| 1977 | 227 |  |  |  |
| 1978 | 91 | 2 |  | 93 |
| 1979 | 235 | 34 |  | 269 |
| 1980 | 203 | 17 |  | 220 |
| 1981 | 1 |  |  | 1 |
| 1982 |  |  |  | 0 |
| 1983 | 65 |  |  | 65 |
| 1984 | 30 |  |  | 30 |
| 1985 | 2 | 6 |  | 8 |
| 1986 | 137 | 57 | 106 | 300 |
| 1987 | 212 | 86 | 92 | 390 |
| 1988 | 91 | 25 | 384 | 500 |
| 1989 | 168 | 8 | 72 | 248 |
| 1990 | 144 |  | 1 | 145 |
| 1991 | 3 |  | 0 | 3 |
| 1992 | 1 |  | 0 | 1 |
| 1993 |  |  | 0 | 0 |
| 1994 | 4 |  | 0 | 4 |
| 1995 | 171 |  | 0 | 171 |
| 1996 | 58 | 2 | 0 | 60 |
| 1997 | 5 |  | 0 | 5 |
| 1998 | 5 |  | 0 | 5 |
| 1999 | - | - | 0 | - |
| 2000 | - | - | 0 | - |
| 2001 | - | - | 0 | - |
| 2002 | - | - | 0 | - |
| 2003 | - | - | 0 | - |

Table 7.1 Nominal shrimp catches ( t ) by country (Sub-areas I and II combined). Data were provided by ICES and Working Group members.

| Year | Norway | Russia | Others | Total |
| ---: | ---: | ---: | ---: | ---: |
| 1970 | 5508 | 0 | 0 | 5508 |
| 1971 | 5116 | 0 | 0 | 5116 |
| 1972 | 6772 | 0 | 0 | 6772 |
| 1973 | 6921 | 0 | 0 | 6921 |
| 1974 | 8008 | 0992 | 0 | 9000 |
| 1975 | 8197 | 0 | 2 | 8199 |
| 1976 | 9752 | 0548 | 0 | 10300 |
| 1977 | 6780 | 12774 | 4854 | 24408 |
| 1978 | 20484 | 15859 | 0 | 36343 |
| 1979 | 25435 | 10864 | 390 | 36689 |
| 1980 | 35061 | 11219 | 0 | 46280 |
| 1981 | 32713 | 10897 | 1011 | 44621 |
| 1982 | 43451 | 15552 | 3835 | 62838 |
| 1983 | 70798 | 29105 | 4903 | 104806 |
| 1984 | 76636 | 43180 | 8246 | 128062 |
| 1985 | 82123 | 32104 | 10262 | 124489 |
| 1986 | 48569 | 10216 | 6538 | 65323 |
| 1987 | 31353 | 6690 | 5324 | 43367 |
| 1988 | 32021 | 12320 | 4348 | 48689 |
| 1989 | 47064 | 12252 | 3432 | 62748 |
| 1990 | 54182 | 20295 | 6687 | 8164 |
| 1991 | 39663 | 29434 | 6156 | 75253 |
| 1992 | 39657 | 20944 | 8021 | 68622 |
| 1993 | 32663 | 22397 | 806 | 55866 |
| 1994 | 20116 | 7108 | 1063 | 28287 |
| 1995 | 19337 | 3564 | 2319 | 25220 |
| 1996 | 25445 | 5747 | 3320 | 34512 |
| 1997 | 29079 | 1493 | 5164 | 35736 |
| 1998 | 44792 | 4895 | $6103^{1}$ | 55790 |
| 1999 | 52612 | 10765 | $12292^{2}$ | 75669 |
| 2000 | 55333 | 19596 | $8241^{3}$ | 83170 |
| 2001 | 43021 | 5875 | $8136^{4}$ | 57032 |
| 2002 | 48799 | 3802 | $8105^{5}$ | 60706 |
| $2003^{6}$ | 34652 | 2775 | $4372^{5}$ | 41800 |
|  |  |  |  |  |

${ }^{1}$ Catches reported by Estonia, Faroe Island, Iceland, Lithuania, Portugal, Spain and UK(Eng.Wal.NI)
${ }^{2}$ Catches reported by Estonia, Faroe Islands, Germany, Greenland, Iceland, Lithuania, Portugal Spain and UK(Eng.Wal.NI)
${ }^{3}$ Catches reported by Estonia, Faroe Islands, Iceland, Lithuania, Portugal, Spain and UK.
${ }^{4}$ Catches reported by Estonia, Faroe Islands, Lithuania, Portugal, Spain and UK
${ }^{5}$ Catches reported by Estonia, Faroe Islands, Lithuania, Spain and UK
${ }^{6}$ Preliminary data

Table 7.2 Catch ( t ), effort ( h ) and CPUE ( $\mathrm{kg} / \mathrm{h}$ ) data in ICES sub-areas I, IIa and IIb. Norwegian data based on $\log$ books from all vessels and scaled to the level of vessels fishing with single trawl at the size of between 1000 hp and 1500 hp . Russian data based on daily reports from vessels smaller than 1300 hp .

|  |  | Norway |  |  | Russia |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | catch (t) | effort | new effort | new CPUE | CPUE | catch (t) | effort | CPUE |
| 1980 | 22822 | 122980 | 95028 | 240 | 186 | - | - | - |
| 1981 | 22458 | 103725 | 80641 | 278 | 217 | 2341 | 8100 | 289 |
| 1982 | 30961 | 156224 | 121602 | 255 | 198 | 4966 | 20400 | 243 |
| 1983 | 54521 | 238724 | 188601 | 289 | 228 | 13223 | 48000 | 276 |
| 1984 | 58489 | 234786 | 187559 | 312 | 249 | 33403 | 118900 | 281 |
| 1985 | 61704 | 259503 | 209538 | 294 | 238 | 27974 | 110900 | 252 |
| 1986 | 35650 | 231269 | 190033 | 188 | 154 | 7912 | 33500 | 236 |
| 1987 | 21354 | 178807 | 147747 | 145 | 119 | 3818 | 23900 | 160 |
| 1988 | 21587 | 192451 | 159265 | 136 | 112 | 9010 | 61600 | 146 |
| 1989 | 36673 | 258188 | 216314 | 170 | 142 | 7928 | 53500 | 148 |
| 1990 | 44626 | 292714 | 249792 | 179 | 152 | 17126 | 94500 | 181 |
| 1991 | 34409 | 201912 | 174577 | 197 | 170 | 15532 | 74100 | 210 |
| 1992 | 36861 | 175262 | 154213 | 239 | 210 | 13025 | 57000 | 229 |
| 1993 | 60508 | 289795 | 236255 | 256 | 209 | 11390 | 60000 | 190 |
| 1994 | 19108 | 104789 | 86636 | 221 | 182 | 4521 | 27500 | 164 |
| 1995 | 15662 | 103585 | 85292 | 184 | 151 | 3347 | 26100 | 128 |
| 1996 | 20343 | 106343 | 91242 | 223 | 191 | 5680 | 35300 | 161 |
| 1997 | 24525 | 106406 | 103776 | 236 | 230 | 1507 | 7600 | 198 |
| 1998 | 34080 | 111651 | 112825 | 302 | 305 | 4900 | 21212 | 231 |
| 1999 | 48369 | 165821 | 214860 | 225 | 292 | 6238 | 30900 | 202 |
| 2000 | 51939 | 189028 | 263483 | 197 | 275 | 12204 | 71784 | 170 |
| 2001 | 42158 | 120787 | 190991 | 221 | 349 | 2484 | 16609 | 150 |
| 2002 | 49480 | 121605 | 223785 | 221 | 407 | 3745 | 21773 | 172 |
| 2003 | 33173 | 86163 | 159611 | 208 | 385 | 2775 | 16390 | 127 |
| $2004 *)$ | 14935 | 34599 | 70849 | 211 | 432 | 2077 | 20134 | 102 |

*) Preliminary data

Table 7.3 Indices of shrimp biomass from Norwegian surveys in the years 1982-2002 by main areas.

| Main | A | B | C - Thor | D-Bear | E | F | G | H | Total | Sum. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area | East Finnmark | Tiddly Bank | Iversen Bank | Island Trench | Hopen | Bear Island | Storfjord Trench | Spitsbergen |  | A,B,C, E |
| Strata | 38078 | 6-7 | 10-12 | $\begin{gathered} 5,8,9, \\ 13 \end{gathered}$ | $\begin{gathered} 14-18, \\ 24 \end{gathered}$ | $\begin{aligned} & 19-221 \\ & 31-40 \end{aligned}$ | 41-50 | 51-70 |  |  |
| 1982 | 35 | 34 | 44 | 53 | 66 | 56 | 17 | 22 | 327 | 179 |
| 1983 | 40 | 57 | 61 | 53 | 112 | 52 | 21 | 33 | 429 | 270 |
| 1984 | 40 | 51 | 64 | 60 | 141 | 66 | 20 | 29 | 471 | 296 |
| 1985 | 23 | 17 | 27 | 18 | 96 | 31 | 17 | 17 | 246 | 163 |
| 1986 | 10 | 7 | 13 | 25 | 57 | 34 | 10 | 10 | 166 | 87 |
| 1987 | 29 | 13 | 18 | 23 | 31 | 10 | 9 | 13 | 146 | 91 |
| 1988 | 26 | 18 | 18 | 36 | 32 | 24 | 13 | 14 | 181 | 94 |
| 1989 | 41 | 17 | 13 | 17 | 33 | 53 | 22 | 20 | 216 | 104 |
| 1990 | 31 | 13 | 25 | 42 | 58 | 43 | 27 | 23 | 262 | 127 |
| 1991 | 22 | 28 | 22 | 54 | 120 | 44 | 21 | 10 | 321 | 192 |
| 1992 | 18 | 22 | 33 | 37 | 62 | 38 | 14 | 15 | 239 | 135 |
| 1993 | 17 | 19 | 32 | 29 | 85 | 20 | 12 | 19 | 233 | 153 |
| 1994 | 19 | 8 | 13 | 15 | 52 | 33 | 9 | 12 | 161 | 92 |
| 1995 | 10 | 10 | 11 | 17 | 83 | 33 | 16 | 13 | 193 | 114 |
| 1996 | 21 | 8 | 26 | 26 | 110 | 42 | 21 | 22 | 276 | 165 |
| 1997 | 24 | 34 | 20 | 34 | 116 | 44 | 12 | 16 | 300 | 194 |
| 1998 | 18 | 24 | 41 | 26 | 120 | 72 | 12 | 28 | 341 | 203 |
| 1999 | 17 | 19 | 23 | 21 | 169 | 31 | 21 | 16 | 316 | 227 |
| 2000 | 14 | 29 | 25 | 26 | 102 | 29 | 10 | 12 | 247 | 170 |
| 2001 | 18 | 10 | 30 | 15 | 61 | 25 | 10 | 17 | 184 | 118 |
| 2002 | 11 | 18 | 28 | 16 | 86 | 18 | 9 | 10 | 196 | 143 |
| 2003 | 15 | 17 | 36 | 12 | 94 | 15 | 8 | 16 | 213 | 162 |
| 2004 | 14 | 24 | 22 | 13 | 46 | 14 | 7 | 11 | 151 | 106 |
| \% 03/02 | 34 | -3 | 30 | -22 | 9 | -19 | -12 | 60 | 9 | 14 |
| \% 04/03 | -4 | 38 | -39 | 6 | -51 | -3 | -8 | -33 | -29 | -35 |

Table 7.4 Indices of shrimp biomass (1000 t) from Russian survey in the 1984-2002 by main areas. Catchability of 0.182 is used in the estimate.

| Main <br> Area | A <br> East <br> Finmark | B Tiddly Bank | C-Thor Iversen Bank | E Hopen | $F$ <br> Bear <br> Island | $\mathbf{G}$ <br> Storfiord Trench | H <br> Spits- <br> bergen | I <br> Kola coast |  | Total | Sum. A,B,C,E |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Strata | 1-4 | 6,7,1s | 10-12,25 | 14-18 | $\begin{aligned} & 38-40, \\ & 43-45 \end{aligned}$ | 48-50 | $\begin{array}{\|c\|} \hline 53-55,58- \\ 60, \\ 63-65,58- \\ 70 \end{array}$ | 2s-6s | 7s-8s |  |  |
| 1984 | 38 | 137 | 99 | 254 |  |  |  | 133 |  | 661 | 528 |
| 1985 | 14 | 45 | 74 | 255 |  | 6 | 46 | 19 | 9 | 468 | 388 |
| 1986 | 9 | 19 | 44 | 140 |  | 42 | 127 | 9 | 9 | 399 | 212 |
| 1987 | 16 | 17 | 59 | 107 | 45 | 36 | 27 | 25 | 14 | 346 | 199 |
| 1988 | 14 | 31 | 39 | 49 |  | 22 | 29 | 36 | 13 | 233 | 133 |
| 1989 | 70 | 128 | 57 | 132 | 6 | 60 | 25 | 105 | 20 | 603 | 387 |
| 1990 | 90 | 195 | 119 | 259 | 14 | 110 | 30 | 196 | 15 | 1028 | 663 |
| 1991 | 90 | 153 | 104 | 541 | 9 | 70 | 27 | 155 | 43 | 1192 | 888 |
| 1992 | 80 | 153 | 92 | 409 |  |  |  | 65 | 77 | 876 | 734 |
| 1993 | 45 | 91 | 159 | 382 | 9 |  | 58 | 37 | 111 | 892 | 677 |
| 1994 | 4 | 35 | 48 | 255 | 21 |  |  | 14 | 27 | 404 | 342 |
| 1995 | 5 | 28 | 15 | 80 | 33 | 53 |  | 16 | 18 | 248 | 128 |
| 1996 | 20 | 98 | 127 |  | 21 |  |  | 67 | 108 | 441 | 245 |
| 1997 | 26 | 108 | 130 | 341 |  |  |  | 108 | 52 | 765 | 605 |
| 1998 | 14 | 106 | 136 | 172 |  |  |  | 108 | 41 | 576 | 427 |
| 1999 | 43 | 139 | 107 | 523 |  |  |  | 93 | 61 | 966 | 812 |
| 2000 | 29 | 73 | 109 | 328 | 9 | 39 |  | 72 | 141 | 800 | 539 |
| 2001 | 11 | 52 | 105 | 185 | 19 | 14 | 13 | 14 | 55 | 468 | 353 |
| 2002 | 30 | 129 | 198 | 353 | 15 | 39 | 51 | 70 | 105 | 980 | 710 |
| \% 01/00 | -62 | -29 | -4 | -44 | 111 | -64 |  | -81 | -61 | -42 | -35 |
| \% 02/01 | 173 | 148 | 89 | 91 | -21 | 179 | 292 | 400 | 91 | 109 | 101 |

Table 7.5 Shrimp in the Barents Sea defined as index of numbers in size groups according to carapace length at age and number of egg bearing females contributing to the recruitment (SSN) in the Norwegian Barents sea survey (whole mm ).

| $\mathrm{CL}(\mathrm{~mm})$ <br> year | <9 |  | 9<cl<13 |  | $13<\mathrm{cl}<17$ | 17<cl<19 | >19mm | SSN |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 |  | 2 |  | 3 | 4 | 5+ |  |
| 1990 |  |  |  | 8 | 192 | 357 | 567 | 131 |
| 1991 |  |  |  | 59 | 213 | 391 | 756 | 123 |
| 1992 |  |  |  | 84 | 308 | 291 | 567 | 109 |
| 1993 |  |  |  | 44 | 355 | 316 | 405 | 101 |
| 1994 |  |  |  | 23 | 186 | 221 | 250 | 30 |
| 1995 |  | 0,4 |  | 20 | 238 | 233 | 307 | 9 |
| 1996 |  | 0,2 |  | 27 | 335 | 374 | 367 | 25 |
| 1997 |  | 0,5 |  | 22 | 372 | 511 | 440 | 47 |
| 1998 |  | 0,8 |  | 9 | 374 | 517 | 567 | 51 |
| 1999 |  | 1,3 |  | 12 | 192 | 357 | 510 | 111 |
| 2000 |  | 2,6 |  | 33 | 147 | 278 | 559 | 66 |
| 2001 |  | 2,1 |  | 20 | 138 | 138 | 410 | 61 |
| 2002 |  | 1,1 |  | 22 | 218 | 295 | 390 | 165 |
| 2003 |  | 0,5 |  | 19 | 254 | 249 | 362 | 110 |
| 2004 |  | 0,7 |  | 5 | 106 | 198 | 295 | 75 |

Table 7.6 Biomass indices for shrimp from the Norwegian surveys, biomass estimate for cod (age 3 years and older) and the shrimp consumed by the cod in the Barents Sea.

| Year | Cod (3+) |  |  |  | Shrimp index Shrimp consumed |  |
| ---: | ---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 818 | 471 | 436 |  |  |  |
| 1985 | 957 | 246 | 155 |  |  |  |
| 1986 | 1292 | 166 | 142 |  |  |  |
| 1987 | 1120 | 146 | 191 |  |  |  |
| 1988 | 913 | 181 | 129 |  |  |  |
| 1989 | 891 | 216 | 132 |  |  |  |
| 1990 | 963 | 262 | 194 |  |  |  |
| 1991 | 1560 | 321 | 188 |  |  |  |
| 1992 | 1910 | 239 | 373 |  |  |  |
| 1993 | 2355 | 233 | 315 |  |  |  |
| 1994 | 2149 | 161 | 516 |  |  |  |
| 1995 | 1815 | 193 | 362 |  |  |  |
| 1996 | 1700 | 276 | 341 |  |  |  |
| 1997 | 1526 | 300 | 311 |  |  |  |
| 1998 | 1221 | 341 | 325 |  |  |  |
| 1999 | 1097 | 316 | 256 |  |  |  |
| 2000 | 1108 | 247 | 459 |  |  |  |
| 2001 | 1393 | 184 | 288 |  |  |  |
| 2002 | 1593 | 196 | 247 |  |  |  |
| 2003 | 1815 | 212 | 285 |  |  |  |
| 2004 | 1749 | 151 |  |  |  |  |


| Skagerrak, Sub-div. IIIA. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species: | 1994 |  | 1995 |  | 1996 |  | 1997 |  | 1998 |  | 1999 |  | 2000 |  | 2001 |  | 2002 |  | 2003 |  |
|  | Tons | \% of total catch | Tons | \% of total catch | Tons | \% of total catch | Total | \% of total catch | Total | $\%$ of total catch | Total | \% of total catch | Total | \% of total catch | Total | \% of total catch | Total | \% of total catch | Total | \% of total catch |
| Blue Whiting | 289.2 | 9.5 | 151.6 | 4.8 | 88.5 | 2.0 | 97.5 | 2.3 | 53.4 | 1.5 | 8.1 | 0.5 | 1.4 | 0.1 | 0.1 | 0.0 | 128.4 | 5.2 | 0.0 | 0.0 |
| Norway lobster | 17.3 | 0.6 | 28.3 | 0.9 | 65.5 | 1.5 | 38.8 | 0.9 | 31.0 | 0.9 | 22.1 | 1.3 | 18.6 | 0.8 | 14.4 | 1.0 | 13.9 | 0.6 | 31.2 | 1.0 |
| Pandalus | 2001.0 | 66.0 | 2421.0 | 76.1 | 3664.2 | 82.1 | 3617.0 | 84.4 | 2933.0 | 83.0 | 1398.5 | 81.8 | 1897.6 | 83.9 | 1185.9 | 84.3 | 1966.6 | 79.2 | 2562.2 | 83.0 |
| Angler fish | 16.2 | 0.5 | 12.3 | 0.4 | 28.5 | 0.6 | 18.7 | 0.4 | 12.5 | 0.4 | 8.0 | 0.5 | 12.4 | 0.5 | 10.0 | 0.7 | 13.2 | 0.5 | 6.6 | 0.2 |
| Whiting | 1.1 | 0.0 | 0.1 | 0.0 | 0.9 | 0.0 | 0.9 | 0.0 | 0.2 | 0.0 | 0.4 | 0.0 | 0.4 | 0.0 | 0.0 | 0.0 | 0.3 | 0.0 | 1.0 | 0.0 |
| Haddock | 16.4 | 0.5 | 10.8 | 0.3 | 19.8 | 0.4 | 9.3 | 0.2 | 17.8 | 0.5 | 9.7 | 0.6 | 11.3 | 0.5 | 13.1 | 0.9 | 72.1 | 2.9 | 79.5 | 2.6 |
| Hake | 8.3 | 0.3 | 3.9 | 0.1 | 7.3 | 0.2 | 6.2 | 0.1 | 2.9 | 0.1 | 2.8 | 0.2 | 3.8 | 0.2 | 7.5 | 0.5 | 4.7 | 0.2 | 5.0 | 0.2 |
| Ling | 1.3 | 0.0 | 0.7 | 0.0 | 1.1 | 0.0 | 0.4 | 0.0 | 0.7 | 0.0 | 0.6 | 0.0 | 0.5 | 0.0 | 0.4 | 0.0 | 0.5 | 0.0 | 1.0 | 0.0 |
| Saithe | 7.7 | 0.3 | 6.0 | 0.2 | 82.6 | 1.9 | 80.8 | 1.9 | 85.6 | 2.4 | 41.0 | 2.4 | 53.9 | 2.4 | 52.6 | 3.7 | 129.1 | 5.2 | 210.2 | 6.8 |
| Witch flounder | 19.5 | 0.6 | 39.8 | 1.3 | 32.5 | 0.7 | 33.8 | 0.8 | 66.6 | 1.9 | 56.1 | 3.3 | 104.5 | 4.6 | 32.6 | 2.3 | 37.6 | 1.5 | 42.8 | 1.4 |
| Norway pout | 383.1 | 12.6 | 144.3 | 4.5 | 114.6 | 2.6 | 83.9 | 2.0 | 29.9 | 0.8 | 0.5 | 0.0 | 0.0 | 0.0 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Cod | 100.2 | 3.3 | 159.0 | 5.0 | 179.5 | 4.0 | 187.8 | 4.4 | 188.4 | 5.3 | 100.9 | 5.9 | 84.7 | 3.7 | 51.6 | 3.7 | 72.6 | 2.9 | 87.4 | 2.8 |
| Other market fish | 6.0 | 0.2 | 6.2 | 0.2 | 5.7 | 0.1 | 5.1 | 0.1 | 6.1 | 0.2 | 3.7 | 0.2 | 6.2 | 0.3 | 2.9 | 0.2 | 4.5 | 0.2 | 7.1 | 0.2 |
| Fish for reduction | 164.0 | 5.4 | 196.9 | 6.2 | 173.4 | 3.9 | 106.6 | 2.5 | 105.6 | 3.0 | 57.7 | 3.4 | 65.5 | 2.9 | 35.0 | 2.5 | 40.6 | 1.6 | 53.9 | 1.7 |
| Cod as \% of shrimp: |  | 5.0 |  | 6.6 |  | 4.9 |  | 5.2 |  | 6.4 |  | 7.2 |  | 4.5 |  | 4.4 |  | 3.7 |  | 3.4 |
| Total H.C. as \% of shrimp: |  | 9.7 |  | 11.0 |  | 11.6 |  | 10.6 |  | 14.0 |  | 17.5 |  | 15.6 |  | 15.6 |  | 17.7 |  | 18.4 |



| C: <br> Skagerrak, Sub-div. |  |  |  |  |  | Swedis | Swedis | h log bo | Trawls | with sele | tive gri | ids |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 199 |  | 1995 |  | 1996 |  | 1997 |  | 1998 |  | 1999 |  | 2000 |  | 2001 |  | 2002 |  | 2003 |  |
| Species: | Tons | $\begin{gathered} \% \text { of total } \\ \text { catch } \end{gathered}$ | Tons | $\begin{gathered} \text { \% of totall } \\ \text { catch } \end{gathered}$ | Tons | $\%$ of tota catch | Total | $\begin{gathered} \% \text { of total } \\ \text { catch } \end{gathered}$ | Total | $\begin{gathered} \% \text { of total } \\ \text { catch } \end{gathered}$ | Total | $\begin{gathered} \% \text { of total } \\ \text { catch } \end{gathered}$ | Total | $\%$ of total catch | Total | $\%$ of total catch | Total | \% of total catch | Total | $\begin{gathered} \% \text { of total } \\ \text { catch } \end{gathered}$ |
| Blue Whiting |  |  |  |  |  |  |  | 0.0 |  | 0.0 |  | 0.0 |  |  |  | 0.0 |  | 0.0 | 0.0 | 0.0 |
| Norway lobster |  |  |  |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |  | 0.0 | 0.0 | 1.0 | 0.6 | 0.4 | 0.2 |
| Pandalus |  |  |  |  |  |  | 1.0 | 100.0 | 35.0 | 100.0 | 1.0 | 100.0 | 0.0 |  | 21.0 | 100.0 | 177.0 | 99.4 | 232.7 | 98.5 |
| Angler fish |  |  |  |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Whiting |  |  |  |  |  |  |  | 0.0 |  | 0.0 |  | 0.0 |  |  |  | 0.0 |  | 0.0 | 0.1 | 0.0 |
| Haddock |  |  |  |  |  |  |  | 0.0 |  | 0.0 |  | 0.0 |  |  |  | 0.0 |  | 0.0 | 0.4 | 0.2 |
| Hake |  |  |  |  |  |  |  | 0.0 |  | 0.0 |  | 0.0 |  |  |  | 0.0 |  | 0.0 | 0.0 | 0.0 |
| Ling |  |  |  |  |  |  |  | 0.0 |  | 0.0 |  | 0.0 |  |  |  | 0.0 |  | 0.0 | 0.0 | 0.0 |
| Saithe |  |  |  |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |  | 0.0 | 0.0 |  | 0.0 | 2.0 | 0.8 |
| Witch flounder |  |  |  |  |  |  |  | 0.0 |  | 0.0 |  | 0.0 |  |  |  | 0.0 |  | 0.0 | 0.2 | 0.1 |
| Norway pout |  |  |  |  |  |  |  | 0.0 |  | 0.0 |  | 0.0 |  |  |  | 0.0 |  | 0.0 | 0.0 | 0.0 |
| Cod |  |  |  |  |  |  |  | 0.0 |  | 0.0 |  | 0.0 |  |  |  | 0.0 |  | 0.0 | 0.2 | 0.1 |
| Other market fish |  |  |  |  |  |  |  | 0.0 |  | 0.0 |  | 0.0 |  |  |  | 0.0 |  | 0.0 | 0.2 | 0.1 |
| Fish for reduction |  |  |  |  |  |  |  | 0.0 |  | 0.0 |  | 0.0 |  |  |  | 0.0 |  | 0.0 | 0.0 | 0.0 |
| Cod as \% of shrimp: <br> Total H.C. as \% of shimp: |  |  |  |  |  |  |  | 0.0 |  | 0.0 0.0 |  | 0.0 |  |  |  | 0.0 |  | 0.0 0.6 | 236.2 | 0.1 1.5 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| D: Skagerrak, Sub-div. |  |  |  |  |  | Norweg | ian log | book reco | ords (* $n$ | ew log b | ok form | mat) |  |  |  |  |  |  |  |  |
|  | 199 |  | 1995 |  | 1996 |  | 199 |  | 1998 |  | 1999 |  | 2000 |  | 2001 |  | 2002* |  | 2003* |  |
| Species: | Tons | \% of total catch | Tons | $\% \text { of total }$ catch | Tons | $\%$ of tota catch | Total | $\% \text { of total } \\|$ catch | Total | $\%$ of total catch | Total | $\%$ of total catch | Total | $\%$ of total catch | Total | $\%$ of total catch | Total | \% of total catch | Total | $\%$ of tota catch |
| Blue Whiting |  |  |  |  |  |  |  |  | 12.0 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Norway lobster |  |  |  |  |  |  |  |  | 3.0 | 0.2 | 7.0 | 0.5 | 9.0 | 0.8 | 20.0 | 1.2 | 37.0 | 0.9 | 28.0 | 0.7 |
| Pandalus |  |  |  |  |  |  |  |  | 1689.0 | 87.5 | 1328.0 | 87.9 | 1031.0 | 86.2 | 1461.0 | 88.3 | 3663.0 | 87.3 | 3700.0 | 86.3 |
| Angler fish |  |  |  |  |  |  |  |  | 9.0 | 0.5 | 11.0 | 0.7 | 13.0 | 1.1 | 13.0 | 0.8 | 32.0 | 0.8 | 26.0 | 0.6 |
| Whiting |  |  |  |  |  |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 12.0 | 0.3 | 14.0 | 0.3 |
| Haddock |  |  |  |  |  |  |  |  | 1.0 | 0.1 | 4.0 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Hake |  |  |  |  |  |  |  |  | 1.0 | 0.1 | 1.0 | 0.1 | 2.0 | 0.2 | 2.0 | 0.1 | 6.0 | 0.1 | 6.0 | 0.1 |
| Ling |  |  |  |  |  |  |  |  | 4.0 | 0.2 | 5.0 | 0.3 | 6.0 | 0.5 | 4.0 | 0.2 | 26.0 | 0.6 | 28.0 | 0.7 |
| Saithe |  |  |  |  |  |  |  |  | 15.0 | 0.8 | 27.0 | 1.8 | 26.0 | 2.2 | 34.0 | 2.1 | 43.0 | 1.0 | 58.0 | 1.4 |
| Witch flounder |  |  |  |  |  |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 39.0 | 0.9 | 34.0 | 0.8 |
| Norway pout |  |  |  |  |  |  |  |  | 41.0 | 2.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Cod |  |  |  |  |  |  |  |  | 30.0 | 1.6 | 25.0 | 1.7 | 24.0 | 2.0 | 20.0 | 1.2 | 153.0 | 3.6 | 184.0 | 4.3 |
| Other market fish |  |  |  |  |  |  |  |  | 126.0 | 6.5 | 103.0 | 6.8 | 85.0 | 7.1 | 101.0 | 6.1 | 187.0 | 4.5 | 208.0 | 4.9 |
| Fish for reduction |  |  |  |  |  |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Cod as \% of shrimp: Total H.C. as \% of shrimp: |  |  |  |  |  |  |  |  |  | $\begin{array}{r} 1.8 \\ 13.6 \\ \hline \end{array}$ |  | $\begin{array}{r} 1.9 \\ 13.8 \\ \hline \end{array}$ |  | $\begin{array}{r} \hline 2.3 \\ \hline 16.0 \\ \hline \end{array}$ |  | $\begin{array}{r} 1.4 \\ \hline 13.3 \end{array}$ |  | $\begin{array}{r} \hline 4.2 \\ \hline 14.6 \\ \hline \end{array}$ |  | 5.0 15.8 |



|  | iv. IVA. |  |  |  | anish | og book | records |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1994 |  | 1995 |  | 1996 |  | 1997 |  | 1998 |  | 1999 |  | 2000 |  | 2001 |  | 2002 |  | 2003 |  |
| Species: | Tons | \% of total catch | Tons | $\begin{gathered} \% \text { of total } \\ \text { catch } \end{gathered}$ | Tons | \% of total catch | Total | $\%$ of total catch | Total | $\begin{gathered} \text { \% of total\| } \\ \text { catch } \end{gathered}$ | Total | $\%$ of total catch | Total | $\%$ of tota catch | Total | \% of total catch | Total | $\begin{gathered} \text { \% of total } \\ \text { catch } \end{gathered}$ | Total | $\begin{aligned} & \text { ff total } \\ & \text { atch } \end{aligned}$ |
| Blue Whiting | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Norway lobster | 31.7 | 2.2 | 119.2 | 2.2 | 104.2 | 2.2 | 44.0 | 1.3 | 134.6 | 3.9 | 40.2 | 3.2 | 42.5 | 2.5 | 17.6 | 1.3 | 37.8 | 2.7 | 21.7 | 1.9 |
| Pandalus | 1229.1 | 86.1 | 4658.5 | 85.5 | 3858.4 | 82.6 | 3022.2 | 89.0 | 2899.8 | 84.1 | 1004.6 | 80.5 | 1482.4 | 86.6 | 1263.3 | 92.5 | 1147.1 | 81.9 | 1001.3 | 85.6 |
| Angler fish | 37.5 | 2.6 | 145.3 | 2.7 | 192.5 | 4.1 | 60.1 | 1.8 | 57.9 | 1.7 | 28.2 | 2.3 | 30.5 | 1.8 | 19.0 | 1.4 | 28. | 2.0 | 19.9 | 1.7 |
| Whiting | 9.4 | 0.7 | 9.3 | 0.2 | 6.0 | 0.1 | 0.6 | 0.0 | 2.1 | 0.1 | 0.5 | 0.0 | 2.5 | 0.1 | 0.2 | 0.0 | 2.7 | 0.2 | 0.6 | 0.0 |
| Haddock | 17.5 | 1.2 | 54.0 | 1.0 | 59.3 | 1.3 | 16.2 | 0.5 | 34.8 | 1.0 | 49.7 | 4.0 | 33.4 | 2.0 | 4.1 | 0.3 | 20.0 | 1.4 | 28.5 | 2.4 |
| Hake | 0.0 | 0.0 | 0.3 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 |
| Ling | 1.9 | 0.1 | 6.2 | 0.1 | 3.2 | 0.1 | 1.1 | 0.0 | 1.0 | 0.0 | 0.4 | 0.0 | 0.8 | 0.0 | 0.1 | 0.0 | 0.5 | 0.0 | 0.2 | 0.0 |
| Saithe | 6.4 | 0.4 | 31.9 | 0.6 | 31.9 | 0.7 | 9.7 | 0.3 | 50.2 | 1.5 | 27.4 | 2.2 | 21.0 | 1.2 | 19.3 | 1.4 | 62.2 | 4.4 | 43.0 | 3.7 |
| Witch flounder | 0.2 | 0.0 | 1.2 | 0.0 | 4.1 | 0.1 | 0.4 | 0.0 | 1.0 | 0.0 | 0.2 | 0.0 | 0.8 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 1.7 | 0.1 |
| Norway pout | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Cod | 80.9 | 5.7 | 378.5 | 6.9 | 371.9 | 8.0 | 223.5 | 6.6 | 235.3 | 6.8 | 85.6 | 6.9 | 84.8 | 5.0 | 34.3 | 2.5 | 93.9 | 6.7 | 47.5 | 4.1 |
| Other market fish | 0.6 | 0.0 | 3.3 | 0.1 | 6.2 | 0.1 | 0.2 | 0.0 | 0.5 | 0.0 | 0.1 | 0.0 | 0.5 | 0.0 | 0.0 | 0.0 | 1.0 | 0.1 | 0.2 | 0.0 |
| Fish for reduction | 11.8 | 0.8 | 39.8 | 0.7 | 34.9 | 0.7 | 17.6 | 0.5 | 30.7 | 0.9 | 10.8 | 0.9 | 11.7 | 0.7 | 7.2 | 0.5 | 7.1 | 0.5 | 5.5 | 0.5 |
| Cod as \% of shrimp: |  | 6.6 |  | ${ }^{8.1}$ |  | 9.6 |  | 7.4 |  | 8.1 |  | 8.5 |  | 5.7 |  | 2.7 |  | 8.2 |  | 4.7 |
| Total H.C. as \% of shrimp: |  | 15.1 |  | 16.1 |  | 20.2 |  | 11.8 |  | 17.8 |  | 23.1 |  | 14.6 |  | 7.5 |  | 21.5 |  | 16.3 |

Table 9.1 Northeast Arctic cod weight (tonnes) and numbers taken as bycatch in the shrimp fishery (1983-2002).

| Cod by-catch | 2001 |  | 2002 |  |
| :---: | :---: | :---: | :---: | :---: |
| Fish length (cm) | weight | numbers | weight | numbers |
| 4 | 0 |  |  |  |
| 5 | 0 |  |  |  |
| 6 | 0,06 | 39795 |  |  |
| 7 | 0,12 | 47175 | 3 | 42927 |
| 8 | 1,49 | 379074 | 5 | 190416 |
| 9 | 2,21 | 393002 | 10 | 713464 |
| 10 | 4,75 | 612643 | 15 | 890682 |
| 11 | 15,46 | 1490688 | 16 | 542936 |
| 12 | 28,96 | 2141669 | 22 | 589909 |
| 13 | 44,66 | 2586164 | 27 | 574110 |
| 14 | 44,09 | 2036305 | 34 | 572853 |
| 15 | 45,57 | 1705141 | 46 | 733283 |
| 16 | 44,8 | 1376351 | 56 | 734496 |
| 17 | 34,69 | 885744 | 79 | 1034145 |
| 18 | 26,85 | 575795 | 105 | 1270462 |
| 19 | 26,48 | 481498 | 118 | 1174854 |
| 20 | 22,18 | 344742 | 116 | 828780 |
| 21 | 10,59 | 141811 | 108 | 467373 |
| 22 | 5,52 | 64136 | 103 | 220620 |
| 23 | 4,59 | 46565 | 110 | 138318 |
| 24 | 0,82 | 7320 | 123 | 119107 |
| 25 | 1,86 | 14620 | 129 | 34477 |
| Total | 366 | 15370239 | 1224 | 10873211 |
| Shrimp c.c. | 46704 | 46704 | 42108 | 42108 |
| Cod as \% of shrimp | 0,008 |  | 0,007 |  |



Figure 3.1. The distribution of the Pandalus stocks in the North Sea area as defined by the ICES squares.


Figure 3.2 Shrimp distribution in the Barents Sea according to Surveys conducted in the period August-October 2004.

Figure 4.1


Figure 4.2


Figure 4.3 Mean quarterly carapace length (mm) for Pandalus in Div. Illa and IVaEast


Figure 4.4 Norwegian Trawl Survey Area. Strata 1-16 and depth contour lines.


Figure 4.5 Relation between fitted and observed Biomass indices


Figure 4.6 SPP estimated biomass and international LPUE


Figure 4.7 Estimated international effort and eploitation level from SPP



Figure 7.1 Shrimp landings from ICES areas I, IIa and IIb by Norway, Russia and other countries in the period 1970-2003


Figure 7.2 Unstandardised Norwegian CPUE (N CPUE), standardised CPUE to vessels with 1000-1550hp and single trawl (N new CPUE) and Russian CPUE (R CPUE) for ICES areas I, IIa and IIb.


Figure 7.3 Length distribution in Norwegian shrimp catches in 2000 to 2003.


Figure 7.4 Survey strata are combined to 9 larger areas marked with letters A to K. East Finnmark (A), Tiddly Bank (B), Thor Iversen Bank (C), Hopen (E), Bear Island (F), Storfjord Trench (G), Spitsbergen (H), Kola coast (I) and the Goose Bank (K) (Anon., 2003a).


Figure 7.5 Shrimp biomass indices from Norwegian and Russian surveys in the Barents Sea and Spitsbergen area in 1982-2004.


Figure 7.6 Index for one and three year old shrimp in the Norwegian Survey.


Figure 7.7 Number of individuals aged five or more (5+) and number of egg carrying females (spawning stock number SSn) in the Norwegian Barents Sea survey.


Figure 7.8 Biomass indices from the Norwegian surveys, biomass estimate for cod (age 3 years and older) and the shrimp consumed by the cod in the Barents Sea.

# Annex 1 - On the minimal allowable size of the Barent Sea northern shrimp 

Working Document \# to the Pandalus Assessment Working Group<br>Copenhagen, 27 October - 04 November 2004

# On the minimal allowable size of the Barents Sea northern shrimp 

by

S. V. Bakanev and B. I. Berenboim<br>PINRO, Murmansk, Russia


#### Abstract

Introduction of the total minimal allowable size for the northern shrimp Pandalus borealis is not evident and effective, because of non-uniformity of the length-sex composition of the it population in some areas of the Barents Sea and Spitsbergen, as well as reliable differences in the growth rate, the spawning stock size and abundance of young shrimp.


## Introduction

The international management of fisheries and assessment of the northern shrimp Pandalus borealis stock in the Barents Sea and adjacent waters is carried out predominantly with the use of descriptive summaries of catch statistics, population structure, abundance indexes and expert estimations. This is caused by the lack and low quality of information necessary for the use of analytical stock assessment methods.

There are no yet common standard methods and schemes of development and adoption of decisions on managing the fishery for northern shrimp. The evident lack of fishery statistic information makes also difficult the realization of the precautionary approach to the management of this species stock in the Barents Sea. Management and assessment of most of northern shrimp stocks in different Atlantic areas are also connected with similar problems (Perry et al., 1999; Hvingel and Kingsley, 2000; Koeller et al., 2000; Anon., 2003).

Nevertheless, studying the world experience of fishery for northern shrimp, three main management strategies peculiar for shellfish fishery and other marine organisms can be outlined. Somehow or other, these strategies were tested on the Barents Sea shrimp stock:

1) Limitation of a catch by size/sex;
2) A total allowable catch (TAC);
3) Limitation of fishing efforts.

Realization of a strategy of size/sex limitation is connected with introduction of various protection measures: regulation of fishing gear and techniques, assignment of size limits, minimal mesh sizes in fishing gear and allowable
by-catch of marine organisms of illegal size. In some areas of the Barents Sea the limitation of allowable catch is applied. Direct control of the fishing effect consists as a rule in limitation of total fishing effort (a number of fishing days), as well as in introduction of a prohibition of fishery in some areas because of by-catches of young fish or small shrimp. Establishment of by-catch limits indirectly influences the re-distribution of fishing efforts during shrimp fishery.

Grounded choice of the stock regulatory strategy requires a certain scientific information (Table 1). Lack of such information conditions for the application of formal methods of stock status assessment (for example, summaries of catch statistics, methods of time series extrapolation, a method of analogies, "traffic light" method) and makes difficult choosing of criteria and biological reference points.

Table 1. Scientific information needed by each regulatory strategy of northern shrimp fishery management (by Perry et al., 1999)

| Regulatory strategy | Necessary information |
| :---: | :---: |
| Size limits | - age at first spawning <br> - ratio "yield per recruit" <br> - natural mortality and growth rate <br> - selectivity of fishing gear and survival of escaped individuals |
| TAC/quota | - determination of a stock unit <br> - migrations <br> - abundance (by surveys, by analysis of catches per unit effort) <br> - stock distribution and area stratification exploitation rate <br> - rates of growth, mortality and recruitment <br> - catchability |
| Limitation of fishing efforts | - determination of a stock unit <br> - migrations <br> - recruitment <br> - change of catchability with stock size <br> - growth rate <br> - catch and effort |

Various methods of analytical assessment of stock status are approved for most of shrimp fishing grounds, including the Barents Sea. The most popular analyses among cohort models are VPA and LBA (length based analysis); those among production models are modifications of the Stefansson model and those among stochastic models are state-
space framework and Bayesian approach (Anon., 2003; Anon., 2003a; Skuladottir and Sigurjonsson, 2004). Status estimation of the Barents Sea population and most of others with the use of the given methods is coupled with the following difficulties:

1) Impossibility of accurate determination of age; analysis of length frequencies permits to determine age approximately;
2) Short life cycle and significant annual dynamics of biomass suggest high mortality, including that from cod predation; this leads to impossibility of the grounded application of biological reference "yield per recruit" (Y/R);
3) Value and variability of natural mortality are much higher than those of fishing mortality; that makes the usage of traditional cohort models difficult;
4) Estimation of total biomass is difficult because of the wide variation of shrimp aggregations density over wide areas, as well as seasonal and spatial dynamics of vertical migrations;
5) A complex structure of the population, diversity of dwelling conditions and dependence of most of subpopulations on the external recruitment impede the application of uniform measures of regulation over the whole area (Berenboim and Lysy, 1987; Aschan et al., 2004).

In spite of these difficulties, for the most of stocks this or that type of a strategy is implemented, which is grounded both with the use of formal and analytical methods. Practically for all stocks, including the Barents Sea one, a strategy of limitation by length/sex is realized through introduction of a minimal mesh size in fishing gear. Minimal mesh size for the shrimp fishery in the Barents Sea is 35 mm , on the Grand Bank of Newfoundland and Flemish Cap 40 mm , in the Strait of Maine - 44.5 mm (Clark et al., 2000; Anon., 2004).

In the practice of fishery for $P$. borealis in Atlantic waters the minimal length of shrimp as a measure of regulation of fishery was introduced only for the Norwegian Economic Zone (Anon., 2003a). As a result, the Norwegian authorities have been unilaterally closing for many years quite big Spitsbergen areas for the international shrimp fishery because of by-catches of young shrimp ( $<15 \mathrm{~mm}$ by the carapace length in according Norwegian rules) of more than 10 $\%$ in weight in some catches. However, on the rest grounds of fishery for the Barents Sea shrimp, as well as in waters of Iceland, Greenland, the USA, Canada and in the NAFO Regulatory Area, such a measure of fishery regulation is not provided for ( Clark et al., 2000; Anon., 2004; Bowering and Atkinson, 2004; Siegstad and Hvingel, 2004; Skuladottir and Sigurjonsson, 2004).

The $32^{\text {nd }}$ session of the Joint Russian-Norwegian Fisheries Commission in 2003 had decided to charge scientists of both countries to assess biological conditions for establishing the total minimal allowable size (TMAS) for northern shrimp of the Barents Sea. In order to evaluate the expediency of the total minimal allowable size for northern shrimp of the Barents Sea, we consider in the present paper the length-age structure of shrimp concentrations, sizes of females' maturity, sex ratio and relative abundance of young individuals from some areas of the Barents Sea and Spitsbergen.

## Material and Method

Materials used in the paper are those obtained in the Russian trawl surveys of the northern shrimp Pandalus borealis in the Barents Sea and Spitsbergen areas in 1989-2002. To conduct a comparative analysis of parameters of growth and sex ratio of the northern shrimp the surveyed area was divided into areas, as it is done during trawl surveys (Aschan et al., 1993; Aschan et al., 2004). In all surveys, a bottom shrimp trawl of the Russian production was used, which had minimal internal mesh size of 40 mm and small-mesh insertion (space between knots is $10-12 \mathrm{~mm}$ ) and without a sorting grid. Horizontal opening of a trawl is 14.5 m , the vertical one is 5 m . A standard trawling distance equaled to 3 miles was used in surveys; the trawling speed was 2.6-2.7 knots.

Individuals with weight of 1 kg selected accidentally were subjected to a biological analysis, which included the following operations: length measuring and determination of a sex and stages of gonads' maturity (Aschan et al., 1993).

Carapace lengths (CL) for length frequency information were measured from the posterior margin of the eyestalk to the posterior mid dorsal edge of the carapace. Sex of the northern shrimp was determined by the shape of the endopodite of the first pair of pleopods, distribution of sternal spines on the first segment of abdomen and by the presence of roe (Rasmussen, 1953; McCrary, 1971). $50-\%$ maturity ( $\mathrm{L}_{50}$ ) was determined by a method proposed by Skuladottir (1990 and 1998).


Figure 1. Survey strata are combined to 9 larger areas marked with letters A to K.East Finnmark (A), Tiddly Bank (B), Thor Iversen Bank (C), Hopen (E), Bear Island (F), Storfjord Trench (G), Spitsbergen (H), Kola coast (I) and the Goose Bank (K) (Anon., 2003a).

Data on size-sex composition from each station were combined by areas and years. For each area the length frequency distributions were built up, as well as the percentage ratio was determined between young individuals ( $<15$ mm CL ), males, females with sternal spines and females without them. It was assumed during the analysis of size frequencies that each modal group corresponded to the year class. Statistic treatment was carried out in MS Excel with the use of data analysis package. For comparison of mean values, the parametric t-criterion of Student with the confidence interval $\mathrm{P}=95 \%$ was used considering hypothesis on the normal distribution of a sample's data. To reveal differences in the size-sex structure of the northern shrimp population, mean percentages of young individuals and different sex groups were tested by areas for 1989-2002.

## Results

Size distribution of shrimp samples from all surveyed areas is polymodal and asymmetric (Fig. 2). The selected size groups permit to assess more or less accurately the age of individuals in samples. Analysis of size-variation frequencies has shown that during trawl surveys young shrimp occur in catches from the age of 2, as a rule. Sizes of this age group vary in dependence on the area from 10 to 14.5 mm by the carapace length. Shrimp at the age of $1+$ occur as single individuals and is presented mainly in areas adjoining to Spitsbergen. Sizes of $1+$ shrimp do not usually exceed 10 mm , that agrees well with data of the Norwegian colleagues, who used small-mesh bag (Hansen and Aschan, 2000). $1+$ shrimp from catches taken by small-mesh bag are of $6-10 \mathrm{~mm}$ by the carapace length.


Figure 2. Typical length distributions of P. borealis in the Barents Sea (Spitsbergen (A), Hopen (B), Kola Coast (C) and Goose Bank (D), Russian surveys data for 1993-2000, blue line - males, green - primiparous females, red - multiparous females, black - total).

Analysis of the size structure of catches has revealed 6 age groups in the most areas of investigations. Some years, 7 age groups of shrimp were determined visually, and in the Spitsbergen area - 8-9 age groups. The first and last age groups did not occur in catches every year. Nevertheless, they were being revealed quite often that made possible to determine the average sizes for these ages totally by areas for the whole period of observations. Intervals between modas of year classes, i.e. of age groups, constituted on average 2 mm for the northern areas (the areas of the Bear Island and Spitsberegen) and 3-4 mm for the southern areas (the Kola Coast and Thor Iversen and Tiddly Banks). Older age groups have a big transgression of sizes of individuals of adjacent ages and overlap often each other.

Modal sizes of middle age groups varied by both areas and years. The least modal sizes for the whole period of investigations of northern shrimp at the age of 3-6 were observed for areas of the Bear Island and Spitsbergen. Modal sizes of these ages are lower than long-term means by all areas by $0.2-0.4 \mathrm{~mm}$. The largest sizes of shrimp at the age of 3-8 were in the Hopen Area and in the Kola Coast area.

Strengths of year classes also differ by areas. Nevertheless, in all areas 1-2 age groups are observed, which dominate in catches. For example, on the Goose Bank there mainly shrimp with modal length of 16-18 mm (at the age
of 3-4). In some years, year classes with modal size 21-23 mm carapace length (at the age of 5-6) are observed. In the Kola Coast area, three length classes with different abundances during the whole investigated period can be determined. These classes can be conditionally attributed to the age of 3,4 and 5 . In the area of Tiddly and Thor Iversen Banks shrimp were observed with modal sizes $16-19 \mathrm{~mm}$ and $20-21 \mathrm{~mm}$. The similar distribution was registered in the Hopen area. However, a portion of shrimp with modal size of $22-24 \mathrm{~mm}$ increased there in some years. The Hopen area and Tiddly and Thor Iversen Banks are close to each other. Therefore, the distribution of size-variation frequencies of shrimp catches and strength of year classes in some years have a similar character. In areas of the Bear Island, Storfjord Trench and Spitsbergen, year classes of different strengths were observed in catches, among which shrimp dominated with modal sizes of 14-16, 17-19 and 21-23 mm at the age of 3,4 and 5.

Mean lengths of shrimp of all sex groups in the area of Spitsbergen are higher than in the Barents Sea (Fig. 3). Minimal mean sizes were in males and females of the Goose Bank shrimp.


Figure 3. Average carapace length of different sex groups of shrimp $P$. borealis in the Barents Sea and Spitsbergen areas, 1989-2002 (the Russian surveys data).
Analysis of size composition of the northern shrimp caught during the Russian trawl surveys has shown that a by-catch of young shrimp with carapace length less than 15 mm does not exceed on average $3.5 \%$. High percentage of young shrimp by-catch was registered on the Goose Bank, where it constituted on average $10.3 \%$ for the whole period of investigations (Table 2). Therefore, at the minimal allowable size of 15 mm and a $10-\%$ standard of young shrimp by-catch, the fishery for shrimp on the Goose Bank could be closed. The annual by-catch of young shrimp in this area varied within 1.5-29.7 \%. For other areas, this index did not exceed $4.8 \%$. Quite big year-to-year variability of young shrimp by-catch by areas should be mentioned. Quite often the relative abundance of young shrimp can differ from year to year about ten times. Nevertheless, such a character of dynamics of young shrimp abundance does not lead to big changes of abundance of year classes at the age of 3-4, which dominate in catches.

Table 2. Mean proportions between young northern shrimp, males, females and $L_{50}$ in the Barents Sea and Spitsbergen areas by results of Russian shrimp surveys, 1989-2002.

| Area | Young shrimps <br> $(<15 \mathrm{~mm}), \%$ | Males, \% | Primiparous <br> females, \% | Multiparous <br> females, $\%$ | L50, mm |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Spitsbergen | 4.8 | 57.7 | 26.2 | 16.1 | 22.7 |
| Srorfjord Trench | 3.2 | 66.9 | 26.0 | 7.1 | 23.1 |
| Bear Is. | 4.4 | 67.8 | 16.0 | 16.2 | 21.3 |
| Hopen | 4.7 | 67.7 | 15.6 | 16.7 | 21.8 |
| Tidly Bank, Thor <br> Iversen bank | 2.6 | 65.4 | 15.2 | 19.4 | 22.0 |
| Goose bank | 10.3 | 76.4 | 9.5 | 14.1 | 21.2 |
| Kola Coast | 4.3 | 57.5 | 17.5 | 25.0 | 21.6 |

Ratio between males and females in catches is approximately equal and strives for the proportion $2: 1$. The exclusion is the Goose Bank, where the ratio constitutes $4: 1$.

A value of $50-\%$ maturing of the northern shrimp females $\left(L_{50}\right)$ in the Barents Sea varies on average within 21.2-22.7 mm. Shrimp with such sizes can be attributed to the age of 5-8. The spawning in the southern areas takes probably place at earlier age than in the northern areas. Differences in the value of $L_{50}$ by areas and years aren't statistically confident. At the same time, the average values of $L_{50}$ of shrimp caught on the Goose Bank, nearby the Bear Island and in the Kola Coast area are lower than those in the Hopen area and on the Tiddly and Thor Iversen Banks.

The relative abundances of sex groups and young shrimp ( $<15 \mathrm{~mm} \mathrm{CL}$ ) by different areas were objected to the comparative assessment with the use of parametric $t$-criterion of Student. It was assumed that at the null hypothesis the difference between long-term percentage of sex groups and young individuals by areas is equal to null, and differences between samples are randomicity. In most cases the hypothesis was rejected for the Kola Coast Area and for the Goose Bank. Statistically reliable differences in the relative abundance of shrimp males and females with the sternal spines were observed for the Goose Bank area compared to other areas of the Barents Sea. The reliable differences in the relative abundance of young shrimp were also registered on the Goose Bank and in areas of the Storfjord Trench, Tiddly and Thor Iversen Banks and in the Kola Coast area.

## Discussion

The comparative analysis of length-sex characteristics of the Barents Sea shrimp has shown that there are some differences between groups of this species from different areas, which worth being accounted at developing of management strategies.

Length-age characteristics of shrimp differ reliably by areas. In the southern part of the Barents Sea (the Kola Coast area), growth rate of shrimp at the age of 2-4 is 1.5-2 times higher than in the northern part (Spitsbergen and Storfjord Trench). Differences in growth rates of shrimp from different areas of the Barents Sea have already been observed earlier (Rasmussen, 1953; Teigsmark, 1983; Hansen and Aschan, 2000), as well as in the Icelandic waters (Skuladottir, 1999).

In accordance with data of the Russian trawl surveys, the basis of catches in most areas of the Barents Sea consists of individuals longer than $16-17 \mathrm{~mm}$. Nevertheless, the strength of length year classes differs noticeably by areas and for the whole period of observations on average. Large portion of young shrimp (12-16 mm ) is registered in catches in the southern and eastern Barents Sea (the Kola Coast area and Goose Bank); in the northern areas this portion is less.

For the length composition of shrimp catches from the Kola Coast area the polymodal structure is peculiar with well-defined 3-4 year classes (see fig. 2). In this area, both young and older age groups are well presented. As a rule, 12 modas in the range of $18-24 \mathrm{~mm}$ are observed in the length composition of shrimp concentrations in the Hopen area.

Abundance of the same year classes of the northern shrimp is different in the length composition of concentrations from different areas. Sizes of shrimp in catches of adjacent areas have different modal groups very often, as well as different numbers of modas.

Such differences in the length-age characteristics of catches from different areas are connected to our opinion with the following reasons: varying level of sub-populations reproduction, i. e. different degrees of spawning stock recruitment in these areas, different year-to-year variations of intensity of recruitment with recruits from outside and probable different natural mortality, which is mainly related to predation of cod (Berenboim et al., 2000), as well as, probably, to different exploitation intensity of the population in different parts of the sea. Results of fecundity investigations, as well as of other reproduction properties of the Barents Sea northern shrimp, demonstrate different reproduction abilities of sub-populations (Berenboim, 1989; Berenboim and Sheveleva, 1989).

All population of shrimp $P$. borealis nearby the northeastern border of the Atlantic part of the area is a part of a super(meta)population consisted of a complex spatially discontinuous sub-populations, which have different abilities to self-reproduction and connected to each other by larvae drift. Intensity and direction of larvae drift influence also the recruitment of shrimp stock in different areas of the sea (Berenboim, 1982, 1992; Berenboim and Lysy, 1987).

For the area of the Goose Bank compared to other areas, the difference in the ratio between females and males is big. Most part of males there does not reach a female stage, probably, because of the unfavourable hydrological conditions.

Compared to the Goose Bank and the northern areas of the sea (the Hopen area), higher portion of females is observed in catches in the Kola Coast area that can be connected with relatively high growth rate of shrimp there.

For majority of areas, the introduction of a TMAS of 15 mm will not apparently influence positively the stock status and will not be an effective measure of fishery regulation. Attempts to assess the consequences of introduction of such a standard have been made earlier (Berenboim, 1989). In the 1980's, by-catch of young shrimp less than 15 mm CL did not exceed $4.7 \%$. Such a low by-catch and peculiarities of a population structure of the Barents Sea northern shrimp permitted to conclude that at fishery for shrimp by fishing gear with minimal mesh size of 35 mm there is no need to introduce the additional fishery regulation measure like a minimal allowable size and limitation of a portion of small shrimp.

More over, in the case of introduction of the TMAS for the Barents Sea shrimp, the shrimp fishery on the Goose Bank can be in some years prohibited. And, the Goose Bank area, being an dependent sub-population with relatively low portion of spawning females, presents a dwelling area with a low reproduction potential. Therefore, the introduction here of a minimal allowable size of shrimp in order to conserve the spawning stock will have very low effectiveness. Besides, there are reasons to think that introduction of allowable size of 15 mm by the carapace length will provoke the increase of young shrimp discards.

Minimal allowable size, as a conservation measure aimed at protection of the spawning stock, should not be less than average sizes of individuals, which gave a generation at least two times. Northern shrimp should participate in the spawning at least two times: at first, as a male, after that as a female. The application of such approach for the northern shrimp shows that they should be fished beginning from the age of mass transformation of primiparous females into multiparous females. In this case, the total minimal allowable size under the Barents Sea conditions would constitute about 22 mm by the carapace length. However, the usage of such length as minimal allowable size would lead to the loss of more than half of catch.

If to assume that the introduction of a TMAS is aimed at the conservation of recruitment, then specific distribution of the northern shrimp and a character of their fisheries will not permit to feel the effect of introduction of this fisheries measure. It is known that immature young individuals remain mainly in the water body, in the zone inaccessible for the bottom trawl fisheries. In some areas young individuals prefer areas with less depths than adult mature individuals (Aschan, 2000; Hansen and Aschan, 2000). If necessary, such areas can be closed for fisheries by means of special instructions as it is done for example in the shallow waters of the Flemish Cap Bank (Anon., 2004).

To our opinion, at different stock exploitation degree different scenarios of its management may be used including introduction of a minimal allowable size. A rational approach to solving of this problem requires developing and introducing of:

1) Criteria of a rate of stock exploitation or types of biological reference points with different zones of the stock status;
2) Regulation schemes for each level of the sock exploitation.

The historic experience of gradual introduction of different fish conservation measures shows that very often the first regulation act of fishery for fish and shellfish is establishing of a minimal mesh size of the fishing gear and (or)
minimal allowable size (Perry et al., 1999; Sokolov, 2001). Later on, the regulation scheme included various measures on limitation of catch and effort. In the middle of the 1990's, schemes of rational exploitation of water bioresources from the point of view of the precautionary approach were developed. In accordance with idea of the precautionary approach for the long-term successful exploitation of the stock, the regulation measures should correspond to the stock status; i. e. for each level of the stock exploitation the corresponding complex of fish conservation measures should be applied.

At the same time, the fishery for the northern shrimp by various countries is mainly directed to the catch of the largest individuals, since their price in the world market is much higher than that of the small shrimp. Therefore, fishermen do not desire to use during fishery for shrimp the trawls with less mesh size than fishing rules require. On the other hand, a wish to use only large individuals increases the probability of small individuals discards. At the fishery for shrimp in the North Sea (the ICES Areas IV and IIIA) two types of discards are distinguished (Anon., 2004a). The first type of discards consists of individuals of the least sizes separated during the technological processing and not used by the Norwegian and Swedish industries. It is reckoned that shrimp with carapace length less than 15 mm are separated and not delivered to the world market. The second type of discards was recently marked out and connected with quite high difference in price for the large boiled shrimp and shrimp of middle sizes, which are delivered fresh to the port. Recent time, a portion of shrimp of middle size increased in catches, which can constitute together with fractions of small shrimp from 2 to $11 \%$ of the yearly catch by estimations of the ICES Working Group.

By the present moment, there are no data on discards for the Barents Sea area. A problem of the stock exploitation rate is still debatable as well. By some data, the coefficient of the fishing mortality for the recent 20 years varied within $0.06-0.37$ for the age of 2-5 (Berenboim et al., 2000). However, calculations of the fishing mortality coefficient are carried out with quite large suppositions and cannot be the absolute fishing indices. It is known that total catch of shrimp in the Barents Sea and adjacent waters is for the recent 2-3 years less than the long-term mean indices for the recent 20 years (Aschan et al., 2004). In the middle of the 1980's the catch exceeded 100 thou. t. Nevertheless, no signs of the stock overexploitation were observed. By results of the Russian and Norwegian surveys, the biomass dynamics corresponded to the 7-8-year cycles, and no adequate response to the fishing effort was recorded.

On the basis of the mentioned above one can conclude that there is no evidence to the necessity of introduction of a minimal allowable size at the reached level of fisheries exploitation of the stock.

## Conclusions

Northern shrimp is one of few water animals' species, for which no minimal allowable sizes were introduced almost in all areas of their fishing. There are some objective positions stipulated by biological and marketing peculiarities of the species, at which the introduction of total minimal allowable size for the Barents Sea northern shrimp is not considered as the obvious measure, even at the intensive exploitation of stocks.

1. Minimal allowable size, as a conservation measure aimed at protection of the spawning stock, should not be less than average sizes of individuals, which gave a generation at least two times: at first, as a male, after that as a female. In this case, the minimal allowable size under the Barents Sea conditions would constitute about 22 mm by the carapace length. However, the usage of such length as minimal allowable size would lead to the loss of more than half of catch.
2. Introduction of the total minimal allowable size for the northern shrimp in the Barents Sea will increase the danger of discards of small shrimp from board of vessels where sorting of catches is used.
3. Since the length-sex composition of the spawning stock of the Barents Sea northern shrimp population is inhomogeneous and its distribution is uneven, it is not expedient to introduce a total minimal allowable size for the whole area of the population.

## References

Anon. 2003. Report of scientific council meeting, Dartmouth, Nova Scotia, Canada, 6-11 November 2003, NAFO SCS Doc. 03/25,: 63 p.

Anon. 2003(a). Report of Arctic Fisheries Working Group. ICES CM 2003/ACFM:22,: 460 p.

Anon. 2004 (a). Report on the Pandalus Assessment Working group. ICES C.M.2004/ACFM:5,: 42 p.

Aschan, M. 2000. Spatial variability in length Frequency Distribution and growth of shrimp (Pandalus borealis) in the Barents Sea. J.Northw. Atl.Fish.Sci., 27: 93-105.

Aschan, M., B. Berenboim, S. Bakanev, and K. Sunnana. 2004. Management of the shrimp fishery (Pandalus borealis) in the Barents sea and Spitsbergen area. Management strategies for commercial marine species in northern ecosystems: Proceedings of the $10^{\text {th }}$ Norwegian-Russian Symposium, IMR-PINRO, Bergen, Norway : 94-103.

Aschan, M., Berenhoim, B., Mukhin, S. and Sunnanå, K. 1993. Results of Norwegian and Russian investigations of shrimp (Pandalus borealis) in the Barents Sea and Svalbard area in 1992. ICES CM 1993/K:9,: 22 p.

Berenboim B.I.1982. Reproduction of the shrimp Pandalus borealis in the Barents Sea. Oceanology 22(1):85-89 (in Russian).

Berenboim, B. 1989. On regulation of shrimp Pandalus borealis fishery in the Barents sea. ICES C.M. 1989/K:13 12 p.

Berenboim B. I. 1992. Northern shrimp (Pandalus borealis) of the Barents Sea (biology and fisheries). Murmansk, PINRO : 136 p. (in Russian).
Berenboim, B.I., Dolgov, A.V., Korzhev, V.A. and Yaragina N.A. 2000. The impact of cod on the dynamics of Barents Sea shrimp (Pandalus borealis) as determined by multispecies models. J. Northw. Atl. Fish. Sci. 27: 1-7.

Berenboim, B.I., A.Yu. Lysy. 1987. Effects of oceanographic factors on populational structure of shrimp (Pandalus borealis) in the Barents Sea and Spitsbergen area. The effects of oceanographic conditions on distribution and population dynamics of commercial fish stocks in the Barents Sea: Proceedings of the third Soviet-Norwegian Symposium, Murmansk, 26-28 May 1986, IMR, Bergen, Norway : 243-250.

Berenboim, B.I., G.K. Sheveleva.1989. Data on the deepwater shrimp (Pandalus borealis) fecundity in the Barents Sea. ICES, C.M. 1989/K:16.: 13 p.

Bowering, W.R. and D.B. Atkinson. 2004. Shrimp stocks in Canadian and NAFO waters. Management strategies for commercial marine species in northern ecosystems: Proceedings of the $10^{\text {th }}$ Norwegian-Russian Symposium, IMRPINRO, Bergen, Norway: 117-118.

Clark, S.H., S.X. Cardin, D.F. Schick, P.J. Diodati, M.P.Armstrong, and D. McCarron. 2000. The Gulf of Maine Northern Shrimp (Pandalus borealis) Fishery: A Review of the Record. J.Northw. Atl.Fish.Sci., 27:193-226.

Hansen, H.O., and M. Aschan. 2000. Growth, size- and age-at-maturity of shrimp, Pandalus borealis, at Svalbard related to environmental parameters. J.Northw. Atl.Fish.Sci., 27:83-92.

Hvingel, C., and M.C.S. Kingsley. 2000. The uncertainty of an assessment procedure for the West Greenland stock of northern shrimp, Pandalus borealis. J.Northw. Atl.Fish.Sci., 27: 183-192.

Koeller, P., Savard, D.G. Parsons, and C. FU. 2000. A precautionary approach to assessment and management of shrimp stocks in the Northwest Atlantic. J.Northw. Atl.Fish.Sci., 27: 235-246.

Perry R. I., C.J. Walters, J.A. Boutillier. 1999. A framework for providing scientific advice for the management of new and developing invertebrate fisheries. Reviews in Fish Biology and Fisheries, 9:125-150.

Rasmussen, B. 1953. On the geographical variation in growth and sexual development of deep-sea prawn (Pandalus borealis Kr.). FiskDir. Skr.Ser. HavUnders., 10: 160 p.

Sigstad, H., C.Hvingel. 2004. Shrimp in Greenland waters. Management strategies for commercial marine species in northern ecosystems: Proceedings of the 10th Norwegian-Russian Symposium, IMR-PINRO, Bergen, Norway, Poster.

Skuladottir, U. MS 1990. Defining stocks of Pandalus borealis off northern Iceland using the maximum length and maturity ogive of females as a measure. ICES Symposium on shellfish life histories and shell fishery models in Moncton, No. 95: 16 p.

Skuladottir, U. 1999. Defining populations of northern shrimp, Pandalus borealis, in Icelandic waters using the maximum length and maturity ogive of females. Rit Fiskideildar, 16: 247-262.

Skuladottir, U. 1998. Size at sexual maturity of female northern shrimp (Pandalus borealis) in the Denmark Strait 198593 and a comparison with the nearest Icelandic shrimp population. J.Northw. Atl.Fish.Sci., 24: 27-37.

Skuladottir, U., J. Sigurjonsson. 2004. Pandalus stocks in Icelandic waters: biology, exploitation and management// Management strategies for commercial marine species in northern ecosystems: Proceedings of the 10th NorwegianRussian Symposium, IMR-PINRO, Bergen, Norway :104-116.

Sokolov K.M. 2001. On biological substantiation of the minimal landing size for cod as a fishery regulation measure. Technical regulations and by-catch criteria in the Barents Sea fishery: Proceedings of the $9^{\text {th }}$ PINRO-IMR Symposium, PINRO Press.:128-136.

Teigsmark G.1983. Populations of the deep-sea shrimp (Pandalus borealis Krøyer) in the Barents
Sea. Fiskedir. Skr. Ser. HavUnders., 17: 377-430.

## Annex 2 - The Spanish NE Arctic Shrimp Fishery in 2003

# The Spanish NE Arctic Shrimp Fishery in 2003 

J. M. Casas<br>Instituto Español de Oceanografía, P.O. Box 1552, Vigo, Spain

In 2003 year, the Spanish fleet targeting for shrimp was composed only by two freezer trawlers HP (1650-2000) that worked from June to December in Svalbard area (ICES Division IIb). The gear used was a simple trawl with sorting grates and a mesh size of $40-45 \mathrm{~mm}$ in the codend. All the hauls were carried out in depths between 200 and 500 meters. Catch and effort data for the whole fleet has been collected from the Spanish Fisheries Administration and adjusted according to the data gathered by the observers on board.

In the last quarter of the year one scientific observer recorded the catches of the most important species, fishing effort (hours) and main biological data of shrimp (oblique carapace length, sex and maturity stage) from samples obtained in one of the two vessels.

The table 1 shows the catches and yields of shrimp by quarter and division as well as the distribution of effort (Number of active units and number of days and hours of activity). Also the catches of the others species (by-catch and discarded) were estimated from the vessel where the scientific observer was present.

The peak of activity and biggest yields (around $210 \mathrm{~kg} / \mathrm{h}$ ) were located mainly at the end of the summer and the beginning of the autumn.

The table 2 and figure 1 show the length distribution of shrimp catches as percentage by quarters.

Table 1. Nominal shrimp catches ( kg ) and by-catch of the main species caught by quarter of the Spanish trawl fleet directed to shrimp fishery in Svalbard area (ICES Division IIb) in 2003. Also the effort and yields of shrimp ( $\mathrm{kg} / \mathrm{h}$ ) are shown.

| SPECIES | 1st | 2nd | 3rd | 4th | Total ${ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| ICES DIVISION IIb |  |  |  |  |  |
| Shrimp |  | 2334 | 242892 | 122879 | 368105 |
| Greenland halibut |  |  |  | 158 | 474 |
| Polar cod |  |  |  | 6971 | 20883 |
| Long rough dab |  |  |  | 572 | 1714 |
| Redfish |  |  |  | 263 | 787 |
| Capelin |  |  |  | 1001 | 2999 |
| Blue whiting |  |  |  | 224 | 672 |
| Cod |  |  |  | 1000 | 2995 |
| Haddock |  |  |  | 294 | 881 |
| Thorny skate |  |  |  | 249 | 745 |
| Number of vessels |  | 1 | 2 | 1 | 2 |
| Fishing days |  | 1 | 59 | 31 | 91 |
| Fishing hours |  | 19 | 1121 | 589 | 1729 |
| CPUE Shrimp (kg/h) |  | 123 | 217 | 209 | 213 |

[^2]Table 2: Length distribution of Artic Shrimp from Spanish catches in ICES Division IIb, 2003.

| Length (mm) | 2nd |  | 3rd |  | 4th |  | Total IIb |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | males | females | males | females | males | females | males | females |
| 8.5 | 874 |  | 90959 |  | 46016 |  | 137849 |  |
| 9.0 | 2706 |  | 281579 |  | 142451 |  | 426735 |  |
| 9.5 | 3615 |  | 376175 |  | 190307 |  | 570097 |  |
| 10.0 | 3439 |  | 357886 |  | 181054 |  | 542379 |  |
| 10.5 | 5912 |  | 615212 |  | 311235 |  | 932359 |  |
| 11.0 | 6205 |  | 645737 |  | 326678 |  | 978620 |  |
| 11.5 | 5852 |  | 609029 |  | 308108 |  | 922989 |  |
| 12.0 | 8589 |  | 893873 |  | 452210 |  | 1354672 |  |
| 12.5 | 7416 |  | 771771 |  | 390439 |  | 1169626 |  |
| 13.0 | 11009 |  | 1145633 |  | 579576 |  | 1736218 |  |
| 13.5 | 11128 |  | 1158088 |  | 585876 |  | 1755092 |  |
| 14.0 | 15145 |  | 1576137 |  | 797368 |  | 2388650 |  |
| 14.5 | 22704 |  | 2362741 |  | 1195310 |  | 3580755 |  |
| 15.0 | 25057 |  | 2607625 |  | 1319197 |  | 3951879 |  |
| 15.5 | 22210 |  | 2311278 |  | 1169275 |  | 3502763 |  |
| 16.0 | 32979 | 271 | 3431977 | 28160 | 1736236 | 14246 | 5201192 | 42677 |
| 16.5 | 35075 |  | 3650128 |  | 1846599 |  | 5531801 |  |
| 17.0 | 38060 |  | 3960818 |  | 2003777 |  | 6002655 |  |
| 17.5 | 32091 |  | 3339591 |  | 1689498 |  | 5061181 |  |
| 18.0 | 26828 | 140 | 2791896 | 14614 | 1412419 | 7393 | 4231143 | 22147 |
| 18.5 | 27805 | 360 | 2893625 | 37427 | 1463884 | 18934 | 4385315 | 56721 |
| 19.0 | 25305 | 1115 | 2633382 | 116061 | 1332227 | 58715 | 3990914 | 175891 |
| 19.5 | 19485 | 2029 | 2027772 | 211120 | 1025849 | 106806 | 3073106 | 319954 |
| 20.0 | 18643 | 2008 | 1940090 | 208986 | 981491 | 105726 | 2940224 | 316721 |
| 20.5 | 16867 | 2976 | 1755267 | 309688 | 887989 | 156671 | 2660123 | 469335 |
| 21.0 | 12788 | 3700 | 1330791 | 385095 | 673247 | 194820 | 2016825 | 583615 |
| 21.5 | 5949 | 4879 | 619132 | 507788 | 313219 | 256890 | 938300 | 769557 |
| 22.0 | 3668 | 7365 | 381728 | 766489 | 193116 | 387767 | 578512 | 1161621 |
| 22.5 | 1645 | 11669 | 171238 | 1214331 | 86629 | 614330 | 259512 | 1840330 |
| 23.0 | 565 | 14873 | 58776 | 1547797 | 29735 | 783030 | 89075 | 2345700 |
| 23.5 | 864 | 11636 | 89905 | 1210938 | 45483 | 612613 | 136252 | 1835188 |
| 24.0 | 293 | 13721 | 30448 | 1427868 | 15404 | 722358 | 46145 | 2163946 |
| 24.5 |  | 7590 |  | 789881 |  | 399601 |  | 1197072 |
| 25.0 |  | 7037 |  | 732352 |  | 370497 |  | 1109886 |
| 25.5 |  | 3015 |  | 313775 |  | 158739 |  | 475529 |
| 26.0 |  | 1180 |  | 122783 |  | 62116 |  | 186078 |
| 26.5 |  | 1885 |  | 196121 |  | 99217 |  | 297223 |
| 27.0 |  | 1221 |  | 127050 |  | 64274 |  | 192545 |
| 27.5 |  | 506 |  | 52684 |  | 26653 |  | 79842 |
| 28.0 |  | 535 |  | 55678 |  | 28168 |  | 84381 |
| 28.5 |  | 404 |  | 42054 |  | 21275 |  | 63734 |
| 29.0 |  | 42 |  | 4370 |  | 2211 |  | 6623 |
| $29.5$ |  | 8 |  | $797$ |  | $403$ |  | $1208$ |
| 30.0 |  | 11 |  | 1170 |  | 592 |  | 1773 |
| Total number | 450771 | 100177 | 46910288 | 10425077 | 23731903 | 5274044 | 71092961 | 15799297 |
| Total catch (kg) | 2334 |  | 242892 |  | 122879 |  | 368105 |  |
| Catches Sampled (kg) <br> $N^{o}$ Samplings <br> $\mathrm{N}^{\mathrm{o}}$. Spec. Sampled |  |  |  |  | $\left\lvert\, \begin{aligned} & 8436 \\ & 15 \\ & 3382 \\ & \hline \end{aligned}\right.$ |  | $\begin{aligned} & 8436 \\ & 15 \\ & 3382 \\ & \hline \end{aligned}$ |  |



Figure 1. Length distribution of shrimp in ICES División IIb from samples obtained in November 2003.

## Appendix 1 - Participants list

## Pandalus Assessment Working Group

ICES, Headquarters, 27 October - 5 November 2004

| NAME | ADDRESS | TELEPHONE | FAX | E-MAIL |
| :--- | :--- | :--- | :--- | :--- |
| Sten Munch- <br> Petersen | Danish Institute for <br> Fishery Research <br> Charlottenlund Slot <br> DK-2920 Charlottenlund <br> Denmark | +4533063390 | +45 <br> 33963333 | smp@dfu.min.dk |
| Ole Eigård | Danish Institute for <br> Fishery Research <br> Charlottenlund Slot <br> DK-2920 Charlottenlund <br> Denmark | +4533963300 | +4533963333 | ore@dfu.min.dk |
| Mats Ulmestrand | Institute of Marine Research <br> Box 4 <br> SE-453 21 Lysekil <br> Sweden | +4652318727 | +4652313977 | mats.ulmestrand@fiskeri <br> verket.se |
| Michaela Aschan | Institute of Marine Research <br> P.O. Box 1870 Nordnes <br> N-5817 Bergen <br> Norway | +4777609742 |  |  |
| Knut Sunnanå | Institute of Marine Research <br> P.O. Box 1870 Nordnes <br> N-5817 Bergen <br> Norway | +4777609732 | $1(709) 7727343$ | $1(709) 7724105$ |
| Dasé Miguel Casas | Inst. Español de Oceanografia <br> Centro Oceanográfico de Vigo <br> Cabo Estay - Canido <br> Apdo 1552 <br> E-36200 Vigo <br> Spain | +34986492111 | +34986492351 | mikel.casas@vi.ieo.es |

## Review report

Regarding the process the benefits of having a joint meeting with the NAFO is stated by the WG to be less than hoped for due to the (unexpected) lack of crucial data. The WG however recommends a continuation of the joint meeting in 2005. It may be noted that the benefits from joint meetings should actually be expected to be even higher when a WG is in a situation of change, being forced to take new approaches due to changed data access. That this turned out not to be the case may be due to the need for the WG to use more time to address the new situation. It might be useful if a future joint meeting was based on a clear understanding of what aspects of the work are expected to benefit from the joint meeting and an agenda developed which ensures that there is interaction on these aspects.

Stock identity issues: the WG recommends data for genetic analysis to be collected from Norwegian surveys. Should this lead to terms of reference for ICES groups or is it an internal note for the Norwegian researchers?

## Stock in IVa east and IIIa

It is indicated that discards practices may be changing towards larger volumes and the WG expresses expectation that data will be available from the ongoing discard sampling programmes. However, it is also indicated elsewhere that few shrimp trips are covered by these programmes and not all fisheries are included. Will such data be forthcoming or should one do more than expect that data will be forthcoming?

The reasoning around the stock status depends presently (in the interim between two survey series) very much on LPUE information. It is a serious cause for concern that the explanation about how LPUE data are generated gives rise to comments from the WG with strong reservations about their representativity. The WG mentions technological changes which would have implications for catchability but the extent of these changes is not known. This is a serious concern when these LPUE data are later used as indicators of stock status.

The main problem with this years assessment is the discontinuity of the Norwegian survey data due to a change in vessel/gear in 2003 and change in timing in 2004. Until a new continuity has been established through intercalibration or through the new series being long enough to judge stock trends the assessment will be entirely dependent on commercial LPUE's which seem to be poorly estimated or influenced by technical changes the extent of which are not known. The WG also expresses concerns about the surplus production model used in recent years, independently of the availability of survey data. The WG does not indicate the direction for a replacement model but it seems that there is a need for considerable intersessional work if this stock is to be subject to analytical assessments again. The WG should make a plan of how to to address the three major issues in the next few years - reestablishment of a survey series, development of commercial LPUE's which are representative and with better known relation to stock trends (or with better descriptors of possible changes in catchability) and development of a new approach to assessments.

The conclusion seems to be that the assessment for this stock has lost its survey series for the time being, that the relation between commercial LPUE data and stock trends is shaky and that the assessment model used in the past suffers some problems which needs to be addressed before a new analytical assessment can be undertaken. This basically indicates that there is very little to go on in terms of advice. The approach taken by the WG by largely relying on LPUE's for the stock and effort trends for exploitation is probably the only possible in the present situation but can only form a basis for a decision whether a precautionary TAC is justified or not.

## Stock on Fladen Ground and Farn deep

No comments

## Barents Sea stock

The WG highlights the existing data problems including the lack of adequate length information from catches and discontinuation of one survey and less effort in another. However, it is also stated that some size related data exist and that the discontinued survey was strongly correlated with the survey which has been maintained. When assessment approaches are developed for this stock in the future it should be investigated what can be done on basis of these existing data. There may be a middle way between the size structured analytical assessment which the WG seems to aim for, requiring full length information, and not trying to use the existing information more analytically.


[^0]:    H. C. Andersens Boulevard 44-46 • DK-1553 Copenhagen V • Denmark

    Telephone $+4533386700 \cdot$ Telefax +4533934215
    www.ices.dk •info@ices.dk

[^1]:    *) previous years' estimates of discards were based on Swedish samples and Norweigian survey data.
    The change in the Norwegian survey gear ion 2003 has made it impossible to use Norwegian data for estimates,
    see Sect. 4.2.2

[^2]:    ${ }^{1}$ Total by-catch estimated from the by-catch observed by scientific observed on board in the last quarter.

