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1 **Effects of increasing $p\text{CO}_2$ on life history traits and feeding of the littoral mysid**

2 *Praunus flexuosus*

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4 Erik Sperfeld^{1,2,3,*}, Anders Mangor-Jensen⁴, Padmini Dalpadado³

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6 ¹ Centre for Ecological and Evolutionary Synthesis (CEES), Department of Biosciences,

7 University of Oslo, P.O. Box 1066 Blindern, N-0316 Oslo, Norway

8 ² Leibniz-Institute of Freshwater Ecology and Inland Fisheries, Dep. Experimental

9 Limnology, Alte Fischerhütte 2, OT Neuglobsow, 16775 Stechlin, Germany

10 ³ Institute of Marine Research, P.O. Box 1870 Nordnes, 5817 Bergen, Norway

11 ⁴ Institute of Marine Research, Austevoll Research Station, 5392 Storebø, Norway

12

13 * Corresponding author: e-mail: eriksperfeld@googlemail.com, Tel : +47 2284 4159,

14 ORCID: 0000-0002-3229-407X

15

16 **Abstract**

17 Mysids, an important food web component in the littoral zone of coastal waters, have been
18 neglected so far in ocean acidification research. Juveniles of the littoral mysid *Praunus*
19 *flexuosus* were exposed in the laboratory to four $p\text{CO}_2$ levels (530, 930, 1200, 1600 μatm)
20 for 5 weeks. In addition, juveniles were provided with two different food levels during the
21 experiment. High $p\text{CO}_2$ did not affect survival, but delayed moulting. Juvenile growth
22 decreased and inter-moult period between the last moulting events increased with
23 increasing $p\text{CO}_2$ at low but not at high food supply, suggesting that high food availability
24 is needed to prevent these negative effects of elevated $p\text{CO}_2$. However, small individual
25 juveniles showed lower feeding rates at high $p\text{CO}_2$ compared to the control after prolonged
26 exposure, suggesting decreased activity likely due to impaired metabolism. The subtle
27 negative effects of elevated $p\text{CO}_2$ on life history traits observed in this study suggests that
28 *P. flexuosus* probably has to adapt to counteract adverse effects of predicted high $p\text{CO}_2$,
29 especially when food is limiting.

30

31 **Key-words:** climate change, ocean acidification, high CO_2 , crustacean zooplankton,
32 mysids, exposure experiments, survival, moulting

33

34 **Introduction**

35 The seawater of the world's oceans is slowly changing its carbonate chemistry due to
36 an increasing uptake of anthropogenically produced CO₂, thereby changing the bicarbonate
37 buffering system and reducing the pH (Caldeira and Wickett 2003; Gattuso and Hansson
38 2011). This phenomenon called 'ocean acidification' (OA) can have adverse effects on
39 marine biota depending on phylogeny (Doney et al. 2009; Kroeker et al. 2013; Wittmann
40 and Pörtner 2013) and thus on ocean ecosystem functioning (Nagelkerken and Connell
41 2015). Heavily calcifying organisms, such as coccolithophores, corals, molluscs, or
42 echinoderms, are often negatively affected by elevated CO₂ concentrations (*p*CO₂) in
43 seawater (Kroeker et al. 2013).

44 Marine crustaceans could be more tolerant to OA than the aforementioned groups
45 (Kroeker et al. 2013; Wittmann and Pörtner 2013), even though negative effects of high
46 *p*CO₂ on life history traits have been reported for certain species (reviewed in Kurihara
47 2008; Whiteley 2011). Not surprisingly, OA research on crustaceans focused mainly on
48 specific taxa or groups, such as copepods (Whiteley 2011; Lewis et al. 2013), due to their
49 abundance and central role in pelagic food webs, or commercially valuable crustaceans
50 such as crabs, shrimps, or lobster (Kurihara et al. 2008; Long et al. 2013; Small et al.
51 2016). However, there is still a lack of studies and taxonomic coverage in crustacean OA
52 research compared to other taxonomic groups, preventing robust general interpretations of
53 the effects of high *p*CO₂ on crustaceans (Kroeker et al. 2013). This calls for studies that
54 investigate potential effects of changing pH/*p*CO₂ on other important, but so far neglected,
55 crustacean groups such as mysids.

56 Crustaceans can adjust physiologically to OA-related changes in seawater carbonate
57 chemistry. Such adjustments are important because increasing environmental *p*CO₂ can
58 cause decreasing haemolymph pH, which would compromise protein function and oxygen
59 supply (Melzner et al. 2009; Whiteley 2011). Thus, crustaceans need to compensate acid-

60 base imbalances of their body fluids to keep physiological processes within functional
61 limits (Melzner et al. 2009). Studies on decapod crustaceans have revealed that acid-base
62 balance is closely associated with iono- and osmo-regulation, because these processes
63 share the same mechanisms (e.g. Truchot 1981; Whiteley et al. 2001). Strong iono- and
64 osmo-regulators are likely better in compensating disruptions in haemolymph pH due to
65 their well-developed ion exchange mechanisms and thus, these species are likely less
66 vulnerable to increasing environmental $p\text{CO}_2$ (Whiteley 2011). Though, continuous
67 adjustments in acid-base balance are metabolically costly, thereby reducing energy
68 availability for other energetically costly processes such as growth and reproduction
69 (Seibel and Walsh 2003; Pörtner and Farrell 2008). However, the increasing energetic
70 demands with increasing $p\text{CO}_2$ can be met by sufficient food acquisition and thus, high
71 food supply may mitigate adverse effects of high $p\text{CO}_2$ on growth, reproduction, or
72 calcification (Thomsen et al. 2013; Towle et al. 2015; Ramajo et al. 2016).

73 Mysids, shrimp-like crustaceans, are common and abundant in coastal areas
74 (Mauchline 1980) and are important food web components that can link benthic and
75 pelagic systems (Lehtiniemi and Nordström 2008). Mysids are omnivorous, feeding on
76 phytoplankton, zooplankton, and detritus (Mauchline 1980; Lehtiniemi and Nordström
77 2008) and are important food items for many fish species (Thiel 1996; Nissling et al.
78 2007). Female mysids are characterized by a brood pouch (marsupium) in which the entire
79 development of the embryo and larvae takes place (Mauchline 1980). A ‘clutch’ of
80 embryos is deposited in the marsupium and larvae develop synchronously until juveniles
81 are released within a short time span (Mauchline 1980).

82 We used *Praunus flexuosus* in laboratory exposure experiments as one of the larger
83 littoral mysid species (adults up to 26 mm). *P. flexuosus* is abundant in North-European
84 coastal waters, especially in the littoral zone among macrophytes (Mauchline 1980), and
85 can tolerate large changes in salinity and temperature (Vlasblom and Elgershuizen 1977;

86 McLusky 1979). The global average in ocean surface pH is predicted to decrease by 0.4 to
87 0.7 for 2100 and 2300, respectively, under high greenhouse gas emission scenarios
88 (Caldeira and Wickett 2003; RCP8.5, IPCC 2013). Besides the average decrease in ocean
89 pH, near shore littoral zones can also experience reduced pH due to riverine freshwater
90 input, upwelling processes or biological activity (Hofmann et al. 2011). The current
91 surface pH of the investigated *P. flexuosus* population (caught in a western Norwegian
92 fjord) ranges seasonally from around 8.05 to 8.25 (Omar et al. 2016), but even higher
93 variability can be expected in shallow inshore waters on a daily, tidal, and seasonal basis
94 (e.g. Feely et al. 2008; Wootton et al. 2008). Thus, *P. flexuosus* may have evolved a well-
95 developed acid-base regulation system to withstand strong natural variations in pH/pCO₂.

96 As far as we are aware, there have been no published studies to date that report effects
97 of changing pH/pCO₂ on mysids. To assess potential effects of near future and extended
98 OA scenarios on mysids, we exposed juvenile *P. flexuosus* to four pCO₂ levels (530, 930,
99 1200, and 1600 μ atm) and associated pH (~8, 7.7, 7.6, 7.5). Freshly released juveniles
100 (cohort hatched in the laboratory from field-caught adult females) were exposed for 5
101 weeks to avoid studying acute stress responses to elevated pCO₂ imposed by short-term
102 exposure. We used juveniles in the experiment to investigate a potentially vulnerable life
103 stage and explored survival, moulting, length, mass, and feeding rates. Additionally,
104 juveniles were kept under two different food regimes (high and low food supply) to
105 investigate whether food availability can help compensate for potentially adverse effects of
106 high pCO₂ exposure. We hypothesise adverse effects of increasing pCO₂ on the measured
107 biological response variables. Further we hypothesise that these potentially adverse effects
108 are stronger under food limitation and that feeding rates are higher under high pCO₂ to
109 satisfy enhanced energetic demands.

110

111 **Materials and methods**

112 *Collection of mysids*

113 Adult mysids of the species *Praunus flexuosus* were collected on 8th and 9th June
114 2013 using a small dip net (opening 25 cm diameter, ca. 1.5 mm mesh size) in the littoral
115 zone at Glesvær (N 60° 12.42', E 5° 3.01') 15 km north west of the Austevoll Research
116 Station of the Institute of Marine Research, Bergen, Norway. The mysids were transported
117 to the Austevoll Research Station and distributed to several tanks (~40 L) containing
118 ambient seawater to allow for acclimation to laboratory conditions (~2 weeks). During
119 acclimation, adult mysids were fed *Artemia salina* nauplii (henceforth *Artemia*) *ad libitum*
120 and commercial shrimp food (EZ-Larva, Zeigler Bros. Inc., USA). Females were identified
121 visually by the marsupium (larvae containing brood pouch). The mature females released
122 juveniles synchronously at the beginning of the acclimation period. After 2 weeks, shortly
123 before the next hatching event was expected, 48 females were gently transferred
124 individually to jars (1 L, clear polycarbonate) containing ~700 mL seawater using a tea
125 strainer. Juveniles that had hatched within a 24 hr period were collected from 20 females
126 and pooled (to avoid effects of potential differences in terms of maternal provisioning).
127 From this pool, juveniles were distributed randomly to the experimental jars (5 juveniles
128 per jar, see below and Fig. 1) using a wide-mouth pipette before the start of the $p\text{CO}_2$
129 exposure experiment. The mysids were kept under dim light in a 16:8 h light:dark cycle.
130

131 *Preparation of different seawater $p\text{CO}_2$ levels*

132 Seawater of different $p\text{CO}_2$ levels was prepared according to Sperfeld et al. (2014). An
133 acidic stock solution (pH~5.8) was produced by releasing CO_2 gas into a seawater tank.
134 The acidic stock solution was mixed with ambient seawater (obtained at 150 m depth,
135 pH~7.94) in three additional mixing tanks to prepare seawater with elevated $p\text{CO}_2$ levels
136 (Fig. 1a). Different amounts of acidic stock solution were added to the mixing tanks using
137 dosage pumps that were controlled by pH-electrodes and pre-set pH transmitters. The pH

138 transmitters were programmed to open magnet valves above pre-set pH values of 7.5, 7.6
139 and 7.75, which allowed flow of acidic stock solution into the mixing tanks. The water of
140 different $p\text{CO}_2$ from the mixing tanks were pumped in closed circuits to four header tanks
141 that were mounted above the exposure tanks/jars used for experiments (Fig. 1a),
142 classifying our experimental set-up as B4 design with clumped segregation according to
143 Cornwall and Hurd (2016). An equal water level in mixing and header tanks was
144 controlled by floatation valves. The circulation between the mixing and header tanks was
145 much higher than the drain from the header to the exposure tanks to ensure stable $p\text{CO}_2$
146 levels (see also Sperfeld *et al.* (2014) for further details about the ‘ocean acidification
147 facility’ at the Austevoll Research Station).

148 The investigated *P. flexuosus* population probably experienced an *in situ* pH of around
149 8.05 to 8.25 depending on season (Omar et al. 2016). To simulate decreases in predicted
150 near future pH of 0.4 for year 2100 and decreases in extended future pH of 0.7 for year
151 2300 (Caldeira and Wickett 2003; RCP8.5, IPCC 2013), mysids were exposed to four
152 different $p\text{CO}_2$ levels: $\sim 530 \mu\text{atm}$ (pH 7.94, ambient seawater), $\sim 930 \mu\text{atm}$ (pH 7.72),
153 $\sim 1200 \mu\text{atm}$ (pH 7.61), and $\sim 1600 \mu\text{atm}$ (pH 7.50) (see Table 1 for detailed carbonate
154 chemistry characteristics).

155

156 ***Experimental set-up and protocol***

157 The experiment started with a total of 240 juvenile mysids (mean \pm SD, total length:
158 4.4 ± 0.24 mm, $n = 28$, individual dry mass: 0.078 ± 0.003 mg, $n = 28$ distributed among 3
159 aluminium micro dishes), i.e. with five individuals per experimental jar (Fig. 1b; 4 $p\text{CO}_2$
160 levels \times 12 jars \times 5 juveniles = 240). Additional jars with juveniles had been kept at
161 ambient and the highest $p\text{CO}_2$ level to conduct short-term feeding trials at the end of the
162 experiment (see below). Juveniles were kept in experimental jars (1 L, clear polycarbonate)
163 with a flow-through of filtered seawater ($\text{O}_2 > 85\%$ saturation) of the different $p\text{CO}_2$ levels

164 at ambient temperature ($\sim 12.5^{\circ}\text{C}$). Twelve experimental jars, with a circular opening at the
165 side (4.5 cm diameter) that was covered by a plankton net of $150\mu\text{m}$ mesh size, were
166 placed into three tanks ($\sim 40\text{ L}$) of a given $p\text{CO}_2$ level (Fig. 1b, see Sperfeld *et al.* 2014 for
167 further details about the set-up). The tanks were constantly supplied with the seawater of
168 different $p\text{CO}_2$ (flow rate $\sim 25\text{ L h}^{-1}$) and served as a surrounding water matrix for the
169 experimental jars to stabilize their temperature ($\sim 12.5^{\circ}\text{C}$). Each jar was constantly supplied
170 with the treatment seawater through a silicone tube (6 mm diameter) with a small pipette
171 tip at the end (inflow). The circular opening at the side that was covered by the plankton
172 net served as outflow into the surrounding matrix water of the tanks. Each experimental jar
173 contained approximately 700 mL of seawater (controlled by the water level in the tanks)
174 with a water exchange of approximately every half hour (flow rate $\sim 1.4\text{ L h}^{-1}$). Water
175 exchange between individual jars within a tank was very unlikely due to the strong
176 unidirectional water flow out of each jar. We considered the 40 L tank as the replication
177 unit (i.e. $n = 3$ per $p\text{CO}_2$ level), because jars within one tank could be more similar to each
178 other in terms of non-treatment variables (e.g. temperature, bacterial communities) than
179 jars in different tanks (Cornwall and Hurd 2016). However, potential non-treatment effects
180 imposed by the mixing or header tanks were very unlikely due to the strong flow through
181 character of our experimental facility, diminishing potential bacteria influence or
182 temperature variability as well as ensuring reliably stable $p\text{CO}_2$ levels throughout the
183 experiment (see also Suppl. Mat., Fig. S1).

184 We investigated the interactive effects of elevated $p\text{CO}_2$ levels and limiting food
185 availability to explore the combined effects of multi-stressor scenarios. Because adverse
186 effects of high $p\text{CO}_2$ exposure may only become apparent under food limitation (Thomsen
187 *et al.* 2013), we offered juveniles two different food supply levels. Half of the jars were
188 supplied with high amounts of *Artemia* (*ad libitum*, high food supply), and the remaining
189 jars were supplied with half of the *Artemia* (limiting amounts, low food supply). This

190 resulted in 2 jars per replicate/tank of both the high and low food supply treatment per
191 $p\text{CO}_2$ level (Fig. 1b). Food supply was also adjusted to juvenile growth; at high food
192 supply, juveniles were fed daily with 120 *Artemia* per juvenile at the beginning, 160
193 *Artemia* per juvenile after 7 days, and 200 *Artemia* per juvenile after 17 days (densities of
194 0.171, 0.229, and 0.286 *Artemia* mL^{-1} , respectively). At low food supply, juveniles were
195 fed daily with half of the amounts, 60 *Artemia* at the beginning, 80 *Artemia* after 7 days,
196 and 100 *Artemia* after 17 days per juvenile (densities of 0.086, 0.114, and 0.143 *Artemia*
197 mL^{-1} , respectively). We used short-term (24 h) enriched *Artemia* nauplii (enriched with
198 Larviva Multigain, France) to provide a highly nutritious food for the growth of juveniles.
199 The juveniles grew well by increasing their length in average 2 fold (15-20 fold increase in
200 mass) within the 5 weeks of the exposure experiment, which is in the range of *P. flexuosus*
201 growth observed in a previous laboratory study (Winkler and Greve 2002). The juveniles
202 reached a length of approximately 7 to 10 mm at the end of the experiment and thus were
203 within good margins before maturation (15 mm at 15°C; Winkler and Greve 2002). Even
204 though the low food treatment did not provide the juveniles with low food *sensu stricto*
205 (i.e. juveniles showed good growth), the results indicated that the low food regime still
206 imposed food limitation compared to the high food regime.

207 Before feeding, experimental jars were inspected daily for dead individuals and moults
208 to determine survival and moulting frequency, respectively. The inter-moult period of
209 juveniles (i.e. the number of days between consecutive moult events) was calculated from
210 the mean moulting time of all individuals in a jar. This was possible as juveniles moulted
211 relatively synchronized during the first 5 moults. Debris on the bottom of the jars was
212 removed regularly using a pipette. The experiment was terminated after 38 days (~5
213 weeks) and the surviving individuals were immediately measured for their total length
214 using a stereo microscope and subsequently transferred to pre-weighted aluminium micro

215 dishes for measuring dry mass using an electronic microbalance (Mettler Toledo UMX2,
216 $\pm 1 \mu\text{g}$) after drying for at least 48 hr at 60°C .

217

218 *Short-term feeding trials*

219 Short-term feeding experiments were conducted on 3 consecutive days at the end of
220 the experiment using juveniles from additional jars that had been kept on high food supply
221 at ambient ($530 \mu\text{atm}$) and the highest ($1600 \mu\text{atm}$) $p\text{CO}_2$ level (Fig. 1b). Individual
222 juveniles (5-6 per $p\text{CO}_2$ level and day, overall $n = 16$ per $p\text{CO}_2$) were transferred into small
223 jars filled with 30 mL seawater of the respective $p\text{CO}_2$ level. 30 *Artemia* were added to
224 each jar (start of feeding, density of one *Artemia* mL^{-1}). After 3-4 hours feeding time,
225 juveniles were removed from the feeding jars and immediately measured for their total
226 length and subsequently transferred to pre-weighted aluminium micro dishes for later dry
227 mass measurements (see above). Remaining *Artemia* (dead and alive) were counted and
228 short-term feeding rates (i.e. number of *Artemia* eaten per individual and hour) were
229 determined. Even though most *Artemia* were dead after the short feeding period (in
230 average $\sim 95\%$), they were still available as food as the juvenile mysids did also feed on
231 non-moving *Artemia* at the bottom of the jars (personal observation). Thus, to avoid
232 overestimation of mysid feeding rates, we counted uneaten dead *Artemia* as surviving prey.
233

234 *Carbonate chemistry analysis*

235 Water samples were taken from the experimental (40 L) tanks frequently (i.e. 17 times
236 during the 38 days of the experiment; see also Suppl. Mat., Fig. S1) and immediately
237 analysed for pH and total alkalinity (A_T). pH (total scale) was measured using the
238 spectrophotometric technique (U-2900 spectrophotometer, Hitachi, Japan) with m-cresol
239 purple as indicator dye (Clayton and Byrne 1993). Samples in cuvettes were temperature
240 controlled to approach *in situ* temperatures. The small deviations between measurement

241 temperature (t1) and *in situ* temperature (t2) were corrected using the following equation
242 (Gieskes 1969):

$$243 \quad \text{pH}_{\text{t2}} (\textit{in situ}) = \text{pH}_{\text{t1}} + 0.0114*(\text{t1} - \text{t2})$$

244 A_T was analysed by potentiometric titration (Dickson et al. 2007) using an alkalinity
245 titrator (Radiometer TIM 840 titration manager, Titralab, Germany). Certified reference
246 material provided by Andrew Dickson (Scripps Institution of Oceanography, San Diego,
247 USA; Batch 114) was used to control for uncertainty in A_T measurements.

248 Salinity and temperature were measured using a conductivity meter (Cond 340i,
249 WTW, Germany). Samples for analyses of silicate and phosphate concentrations were
250 preserved in chloroform and measured spectrophotometrically after molybdenum blue
251 reaction using an Skalar autoanalyser (Grasshoff 1965).

252 Carbonate chemistry parameters were calculated using the program CO2sys, version
253 2.1 in Microsoft Excel (Pierrot et al. 2006), with the standard set of carbonate system
254 equations and constants of Mehrbach et al. (1973) after refit of Dickson and Millero
255 (1987). The input variables A_T , salinity, phosphate, and silicate were used as mean values
256 (Table 1) due to low variability and no shift during the experiment.

257

258 ***Data analyses***

259 All statistical analyses and tests of their assumptions were performed using the
260 statistical software R, version 3.2.5 (R Core Team 2016). To account for the B4 design of
261 our CO₂-manipulation system, mixed effect models with exposure tank number (ET-id) as
262 random effect were used to incorporate some source of variance (potentially caused by
263 non-treatment effects) from tank identity (Cornwall and Hurd 2016).

264 Survival was analysed with mixed effects Cox models (Therneau et al. 2003) using the
265 ‘coxme’ and ‘survival’ package (Therneau 2012). Food regime and $p\text{CO}_2$ level were used
266 as fixed effects, and individual was nested within jar to account for the nested design of

267 multiple individuals per jar. Likelihood ratio tests were used to test for differences among
268 survival curves ($p\text{CO}_2$ used as factorial variable) and Holm corrected P -values were given
269 for multiple comparisons.

270 Linear mixed effects (LME) models were applied to the other response variables using
271 the ‘lme4’ package with the random effect as described above. Food regime and $p\text{CO}_2$
272 level were used as fixed effects and mean values of response variables per jar were used in
273 the models (dry mass was log-transformed to meet assumptions). To test for significance of
274 fixed effects, analysis of variance tables of type III with Satterthwaite approximation for
275 degrees of freedom (df) were used, resulting occasionally in non-integer df. Tukey
276 contrasts following LME models were used for multiple comparisons to identify
277 significant differences among treatments. Differences in feeding rates of juveniles between
278 the two tested $p\text{CO}_2$ levels were analysed using individual mass or length as continuous
279 variable (i.e. as fixed effect covariate to account for size differences) and measurement day
280 as random effect.

281

282 **Results**

283 ***Carbonate chemistry***

284 Mixing of ambient seawater with varying quantity of the highly CO_2 -enriched
285 seawater resulted in clearly distinct and elevated $p\text{CO}_2$ levels of approximately 930, 1200,
286 and 1600 μatm in the exposure tanks (Table 1; Fig. S1). The measured pH-values in the
287 CO_2 -enriched tanks matched well the pH transmitter settings. The $p\text{CO}_2$ of ambient
288 seawater with a calculated average of approximately 530 μatm (Table 1) was higher than
289 the atmospheric equilibrium (i.e. 380 μatm), probably because seawater used in this study
290 originated from deeper depth (150 m).

291

292 ***Survival and moulting***

293 The juveniles showed low mortality during the 5 weeks of the experiment (survival
294 ~70%; Fig. 2). Survival was not significantly different among $p\text{CO}_2$ levels when low and
295 high food treatments were combined as well as when low and high food treatments were
296 analysed separately (Table 2; Fig. S2).

297 The juveniles moulted synchronously 5 times during the exposure experiment with
298 clear breaks between the moulting events (Fig. 3a). Very high synchrony was observed
299 during the first three moults, whereas the moulting interval spread more at the fourth and
300 fifth moult (Fig. 3a). A higher number of moults was observed in the ambient control
301 treatment (530 μatm) compared to the CO_2 -enriched treatments at the third, fourth, and
302 fifth moulting event (Fig. 3a; Fig. S3; LME model with CO_2 -enriched treatments pooled;
303 3rd moult: $F(1,10.7) = 8.75, P = 0.013$; 4th moult: $F(1,11.8) = 6.47, P = 0.026$; 5th moult:
304 $F(1,11.3) = 8.22, P = 0.015$). This resulted in a higher cumulative number of moults in the
305 control treatment at the end of the experiment compared to the CO_2 -enriched treatments
306 (Fig. 3b; $F(3,48) = 7.12, P = 0.0005$). These patterns in moulting were very similar in both
307 food treatments (Fig. S4) and the cumulative number of moults was not affected by food
308 regime (Table 3).

309 The inter-moult period increased with the age of juveniles from ~6 days for the first
310 moulting event to ~8 days between the 4th and 5th moult (Table 4). Inter-moult periods
311 were not significantly different among $p\text{CO}_2$ levels up to the fourth moulting event (Table
312 4). However, the inter-moult period between the fourth and fifth moult increased with
313 increasing $p\text{CO}_2$ by almost a day (Table 3, Table 4), and this increase was driven by a
314 significant increase (~1.2 d) in the low but not in the high food treatment (Table 4).

315

316 ***Growth***

317 The food supply regime and $p\text{CO}_2$ level had an effect on mean total length and mean
318 dry mass of the juveniles (Fig. 4, Table 3). In average, juveniles were larger and heavier at

319 high food compared to low food supply, with the effect having a larger magnitude for dry
320 mass than for total length (increase of ~5% in length and ~20% in mass; Fig. 4). The $p\text{CO}_2$
321 effect is mainly driven by the smaller length (8% decrease) and lower mass (21% decrease)
322 at 1600 μatm compared to 530 μatm in the low food treatment (Tukey contrasts following
323 LME model, total length: $P = 0.0187$; $\log_{10}(\text{dry mass})$: $P = 0.0185$). In fact, total length
324 and individual dry mass decreased with increasing $p\text{CO}_2$ in the low food treatment
325 (continuous $p\text{CO}_2$ as fixed effect in LME model, total length: $F(1,24) = 6.92$, $P = 0.015$;
326 $\log_{10}(\text{dry mass})$: $F(1,24) = 6.46$, $P = 0.018$), whereas there was no change with increasing
327 $p\text{CO}_2$ in the high food treatment (continuous $p\text{CO}_2$ as fixed effect in LME model for total
328 length and $\log_{10}(\text{dry mass})$, $F(1,12) < 0.2$, $P > 0.70$).

329

330 **Feeding rates**

331 Feeding rates of juveniles, measured in short-term trials after 5 weeks of exposure to
332 ambient and high $p\text{CO}_2$, were affected by $p\text{CO}_2$ after accounting for varying size of
333 individuals (Fig. 5; LME model with day as random effect, $p\text{CO}_2$: $F(1,29.2) = 17.44$, $P =$
334 0.0002 , individual dry mass as covariate: $F(1,30.2) = 9.20$, $P = 0.0049$, $\text{mass} \times p\text{CO}_2$:
335 $F(1,29.2) = 8.96$, $P = 0.0056$). Smaller individuals that were exposed for 5 weeks to high
336 $p\text{CO}_2$ showed lower feeding rates compared to similar sized individuals that were exposed
337 to ambient $p\text{CO}_2$ (Fig. 5; see also significant interaction term above). Similar results were
338 observed when using total length for accounting for varying individual size (Fig. S5).

339

340 **Discussion**

341 In our laboratory exposure experiment, using juvenile specimens of the littoral mysid
342 *P. flexuosus*, we observed some subtle direct effects of increasing $p\text{CO}_2$ on the measured
343 biological response variables. These negative effects were mainly caused by the highest
344 applied $p\text{CO}_2$ level (1600 μatm), suggesting that this coastal mysid species will not suffer

345 dramatically from predicted near future changes in $p\text{CO}_2$. However, the observed adverse
346 effects of increasing $p\text{CO}_2$ on moulting frequency and inter-moult period suggest an
347 increased selection pressure on moulting as key trait of crustacean life history as these
348 animals need to moult for successful growth and reproduction.

349 It has been postulated that the natural variability in $p\text{CO}_2$ organisms experience may
350 determine their sensitivity or resilience to future OA conditions (Lewis et al. 2013). We
351 exposed *P. flexuosus* to a wide range in $p\text{CO}_2$ levels, as this mysid species inhabits the
352 littoral zone of coastal waters, which can be subject to recurring large $p\text{CO}_2$ fluctuations
353 both diurnally and seasonally (e.g. Feely et al. 2008; Wootton et al. 2008). Our
354 investigated coastal mysid species may have evolved well-developed acid-base regulation
355 systems to tolerate such natural $p\text{CO}_2$ fluctuations. It has been shown in the laboratory that
356 *P. flexuosus* can tolerate large changes in salinity (Vlasblom and Elgershuizen 1977;
357 McLusky 1979) associated with hyper/hypo-osmotic regulation (McLusky and Heard
358 1971; McLusky 1979). These findings together with the wide distribution of *P. flexuosus*
359 across waters differing strongly in salinity and temperature, from coastal Atlantic to the
360 Baltic Sea (Mauchline 1980), suggests that this mysid species has a high acclimation
361 and/or adaptation potential to varying environmental conditions including changes in
362 $p\text{CO}_2$. However, it remains to be further tested whether this high acclimation and/or
363 adaptation potential is indeed beneficial for long-term changes in $p\text{CO}_2$ as imposed by OA.

364

365 ***Effects of $p\text{CO}_2$ on survival and moulting***

366 The survival of juvenile mysids in our experiment was not affected by $p\text{CO}_2$ during the
367 5 weeks of exposure. This indicates no direct lethal effect of $p\text{CO}_2$ on *P. flexuosus* within
368 the range tested and fits to the growing evidence that crustacean survival in average is not
369 reduced significantly by OA-relevant increases in $p\text{CO}_2$ (Kroeker et al. 2013). However,
370 moulting frequency was adversely affected by $p\text{CO}_2$. Juveniles moulted more often in the

371 ambient control treatment compared to all elevated $p\text{CO}_2$ levels, indicating that even
372 relatively small increases in $p\text{CO}_2$ can have adverse effects on moulting of this early life
373 stage. The reduced moulting in all elevated $p\text{CO}_2$ levels was evident only from the 3rd
374 moulting event onwards. We also observed an increasing inter-moult period with
375 increasing $p\text{CO}_2$ between the 4th and 5th moult, but not between earlier moults. These
376 results suggest that adverse effects on some biological response variables become apparent
377 only after prolonged exposure to elevated $p\text{CO}_2$, or alternatively, that later moulting is
378 more sensitive to elevated $p\text{CO}_2$ than earlier moulting.

379 *P. flexuosus* may have a strong buffering mechanism for maintaining osmotic and
380 ionic balance in the case of changing $p\text{CO}_2$ due to its strong hyper/hypo-osmotic regulation
381 capability (McLusky and Heard, 1971; McLusky, 1979; Whiteley 2011). Though, moulting
382 could be impaired directly by slight changes in the animal's acid-base status at elevated
383 $p\text{CO}_2$, influencing the activity of enzymes involved in the moulting process, as enzymes
384 are maximally active only within a narrow pH range (Harvey and Ferrier 2011). An
385 impairment of moulting involved enzymes due to lower pH could explain the negative
386 effects on moulting frequency and inter-moult period observed in our study. Alternatively,
387 moulting could be negatively affected at elevated $p\text{CO}_2$ due to insufficient energy
388 availability, as moulting is an metabolically demanding process (e.g. Chan et al. 1988).

389

390 *Effect of varying food supply on the responses to elevated $p\text{CO}_2$*

391 The inter-moult period of juveniles observed in our study (6 to 9 days) increased with
392 increasing age in accordance to previously described values for *P. flexuosus* (Winkler and
393 Greve 2002). Inter-moult period of crustaceans is known to vary strongly with temperature
394 (e.g. Buchholz 2003), but can also depend to some extent on other external factors such as
395 varying food supply (Buchholz 1991; Qiu et al. 1997). We kept temperature constant but
396 varied food supply in our experiment and found that later inter-moult periods increased

397 significantly with increasing $p\text{CO}_2$ only within the treatments supplied with low food. A
398 similar pattern was also observed for the size and mass of juveniles at the end of the
399 experiment, i.e. the growth of juveniles decreased with increasing $p\text{CO}_2$ at low food supply
400 but not at high food supply. The results on inter-moult period and growth suggest that
401 higher food supply can be used by the animals to compensate the adverse effects of high
402 $p\text{CO}_2$ exposure. In accordance with this, it has been observed that high food availability
403 can outweigh adverse effects of high $p\text{CO}_2$ exposure in juvenile *Mytilus edulis* (Thomsen
404 et al. 2013). Similarly, the Caribbean coral *Acropora cervicornis* can maintain ambient
405 growth rates at elevated $p\text{CO}_2$ via increased feeding when supplied with sufficient food
406 (Towle et al. 2015).

407 Elevated $p\text{CO}_2$ levels can impose increasing maintenance costs, e.g. caused by acid-
408 base homeostatic regulation, thereby reducing the energy available for other energetically
409 costly processes such as growth (Seibel and Walsh 2003; Pörtner and Farrell 2008). These
410 increasing energetic demands at elevated $p\text{CO}_2$ can be met by sufficient food acquisition
411 (Thomsen et al. 2013; Ramajo et al. 2016), thus masking potentially adverse effects that
412 would become visible only during periods of food scarcity. Organisms in natural
413 environments are often confronted with strong variations in food availability, highlighting
414 the need for studies that vary food supply together with stressors such as elevated $p\text{CO}_2$ as
415 done in the current experiment on juvenile mysids. Unfortunately, not much is known
416 about food limiting conditions of *P. flexuosus* in nature. The range of food availability and
417 diversity in the littoral zone is broad and in a study investigating *P. flexuosus* and other
418 littoral mysids in the Baltic Sea it is assumed that food was well available (Lehtiniemi and
419 Nordström 2008). However, considering seasonal variation in food availability, food
420 limitation of *P. flexuosus* cannot be excluded, especially in the rather oligotrophic western
421 North Atlantic, making some populations potentially more vulnerable to predicted future
422 changes in $p\text{CO}_2$.

423

424 ***Feeding rate responses to high pCO₂ after prolonged exposure***

425 The minor negative effect on juvenile feeding rates observed after 5 weeks of
426 exposure to elevated pCO₂ is in contrast to our hypothesis of higher feeding rates at high
427 pCO₂ to meet increased energetic demands. Increased feeding rates, for instance, have
428 been observed for the copepod *Centropages tenuiremis* after short-term exposure (4 days)
429 to OA-relevant pCO₂ with concomitantly increased respiration rates (Li and Gao 2012).
430 Likewise, feeding rates and metabolism increased for freshly *in situ*-caught Antarctic krill
431 at elevated pCO₂ in a short-term (<48 h) exposure experiment (Saba et al. 2012),
432 suggesting also an immediate response to meet energetic demands. Our feeding trials
433 however, were conducted with juveniles kept for a longer time period in the laboratory (5
434 weeks on high food supply) and animals may not sustain the observed short-term
435 adjustments in feeding and associated metabolic rates over longer time periods. Smaller
436 juvenile mysids showed lower feeding rates at high pCO₂ compared to the control at the
437 end of the experiment, suggesting that the prolonged stress slowed down activity probably
438 by impairing metabolic functions (Pörtner and Farrell 2008; Sokolova et al. 2012). This is
439 supported by a study on juvenile European lobster, *Homarus gammarus*, which showed
440 reduced metabolic rates as well as reduced food consumption after 5 weeks exposure to
441 OA-relevant pCO₂ (Small et al. 2016). Thus, prolonged exposure to elevated pCO₂ could
442 have weakened the capability of some individuals to compensate for increased energetic
443 costs by increased feeding, which would ultimately reduce growth. However, we did not
444 observe reduced growth in the treatment with high food supply, where length and mass at
445 the end of the experiment did not decrease with increasing pCO₂. Also in the juvenile
446 European lobster study, inter-moult growth was not lower in the near future OA-treatment
447 (1100 μ atm) compared to the control (Small et al. 2016). However, with feeding rates
448 measured only at the end of the experiment, we cannot exclude that juvenile mysids

449 compensated for increased energetic demands earlier during exposure. Measuring feeding
450 rates across the whole duration of experiments can reveal if and how long organisms
451 mitigate adverse effects of elevated $p\text{CO}_2$ by increased feeding rates.

452

453 ***Conclusions and suggestions for future studies***

454 In our experiments we tested only for direct effects of increasing $p\text{CO}_2$ on biological
455 response variables, but indirect effects via changes in food quality (Rossoll et al. 2012) or
456 species interactions (Diaz-Pulido et al. 2011; Kroeker et al. 2012) may also affect mysid
457 populations in nature. In the present study, we observed some sub-lethal, but potentially
458 important effects of elevated $p\text{CO}_2$ on life history traits of juvenile mysids, suggesting that
459 *P. flexuosus* probably has to adapt to counteract adverse effects of predicted future changes
460 in $p\text{CO}_2$. However, *P. flexuosus* occurs in seawater of very different environmental
461 conditions (Mauchline 1980) and is also the only non-native marine zooplankton species
462 successfully established on the east coast of North America (Ruiz et al. 2011), suggesting a
463 high adaptation potential of this probably highly phenotypically plastic species to changing
464 environmental conditions. Thus, *P. flexuosus* may not suffer dramatically under future OA
465 conditions, but future studies should explore both phenotypic plasticity and the capacity of
466 *P. flexuosus* to adapt over multiple generations (Reusch 2014; Sunday et al. 2014). This
467 could be done by investigating populations from waters differing in natural $p\text{CO}_2$ exposure
468 or by transgenerational studies that require longer-term exposure experiments spanning at
469 least two generations (e.g. Thor and Dupont 2015). Additional studies should also explore
470 the role of $p\text{CO}_2$ variability, because coastal waters, which this littoral species inhabits, can
471 be subject to recurring large fluctuations in $p\text{CO}_2$ both diurnally and seasonally (e.g. Feely
472 et al. 2008; Wootton et al. 2008). We recommend that the $p\text{CO}_2$ range applied in
473 experimental studies should exceed the current natural $p\text{CO}_2$ variation of waters inhabited
474 by *P. flexuosus*. Future studies using mysids should also concomitantly investigate

475 metabolic rates and acid-base status to reveal their capabilities in adjusting physiologically
476 to both short- and long-term changes in $p\text{CO}_2$. A growing number of studies, including this
477 one, indicate the importance of food supply in physiological responses and acclimation to
478 elevated $p\text{CO}_2$, suggesting that variations in food availability should be included in
479 predictions of organisms' responses to future OA scenarios.

480

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488

489 **Compliance with ethical standards**

490 All authors declare they have no conflict of interest. All applicable international,
491 national, and/or institutional guidelines for the care and use of animals were followed.

492

493 **Electronic supplementary material**

494 Figure S1. Measured pH and corresponding $p\text{CO}_2$ during the course of the experiment.

495 Figure S2. Survival curves of juvenile *P. flexuosus* exposed to different $p\text{CO}_2$ levels at both
496 low and high food supply.

497 Figure S3. Average number of moults of juvenile *P. flexuosus* for the five moulting events
498 observed in the experiment.

499 Figure S4. Absolute and cumulative moult number of juvenile *P. flexuosus* over time in the
500 treatments of low and high food supply.

501 Figure S5. Short-term feeding rates (number of *Artemia salina* nauplii eaten per hour)
502 depending on the total length of individual juvenile *P. flexuosus* after exposure of 5
503 weeks to ambient $p\text{CO}_2$ (530 μatm) or 1600 μatm .

504

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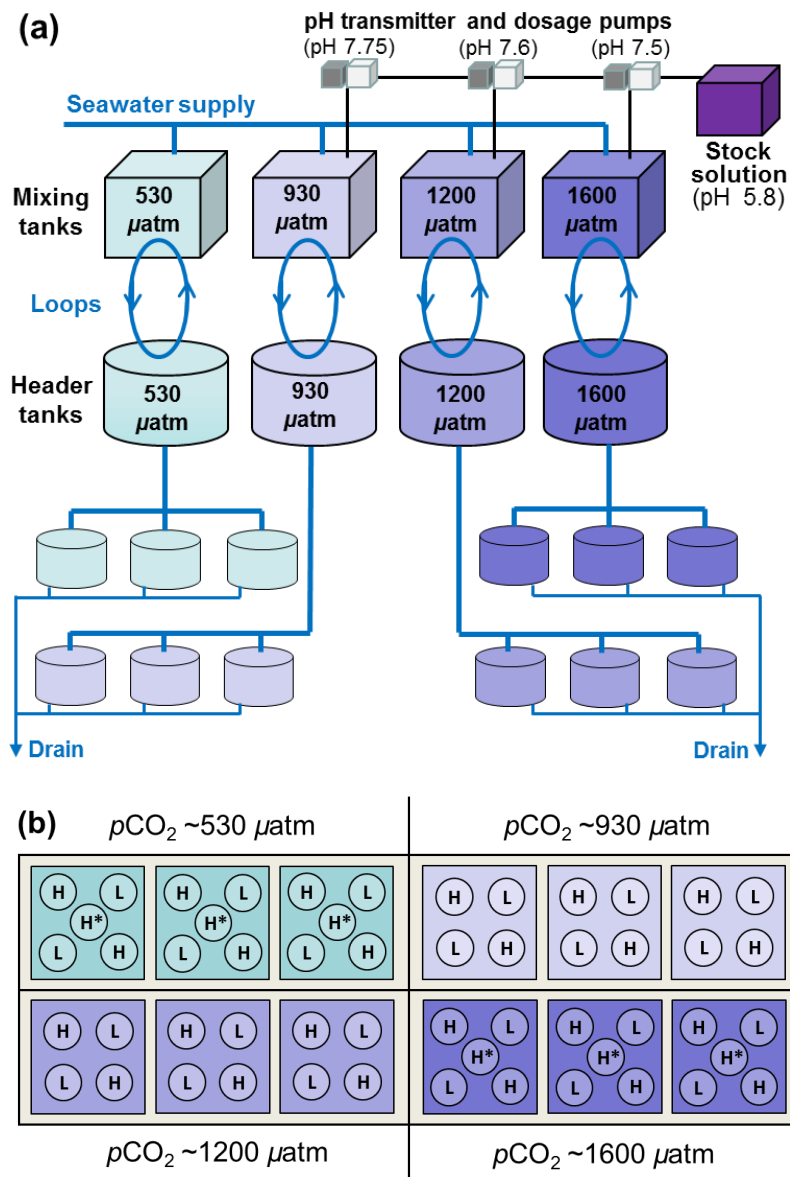


Fig 1 Schematic of (a) the CO_2 manipulation set-up and (b) detailed experimental set-up used in the exposure experiment. Juvenile *Praunus flexuosus* have been exposed to 4 different $p\text{CO}_2$ levels (530, 930, 1200, 1600 μatm) at constant temperature ($\sim 12.5^\circ\text{C}$) in a flow-through system for 5 weeks. The experiment started with 5 individuals per experimental jar (indicated by small circles). Mysids in half of the jars were supplied with high food (H) or low food (L), resulting in 2 jars of both the high and low food treatment per exposure tank as the replication unit per $p\text{CO}_2$ level (i.e. $n=3$). Asterisks indicate additional jars containing juvenile mysids that have been used in feeding trials after 5 weeks exposure. Experimental jars were placed in tanks providing a surrounding water matrix with the same $p\text{CO}_2$ level (indicated by quadrats, three tanks per $p\text{CO}_2$ level) and there was a unidirectional flow of water through the jars into the surrounding water matrix.

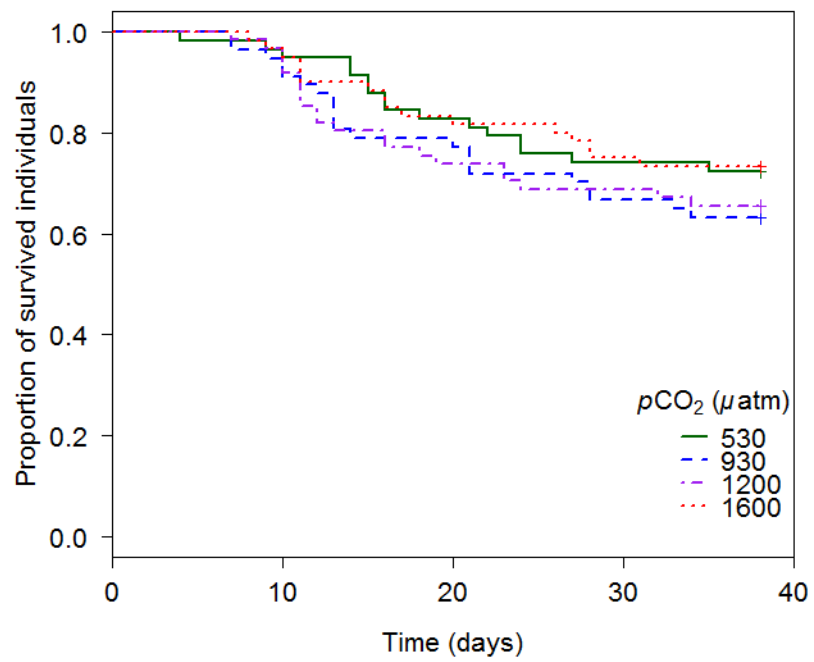


Fig 2 Survival curves of juvenile *P. flexuosus* exposed to different $p\text{CO}_2$ levels (low and high food combined, see Fig. S2 for graphs of both low and high food treatments).

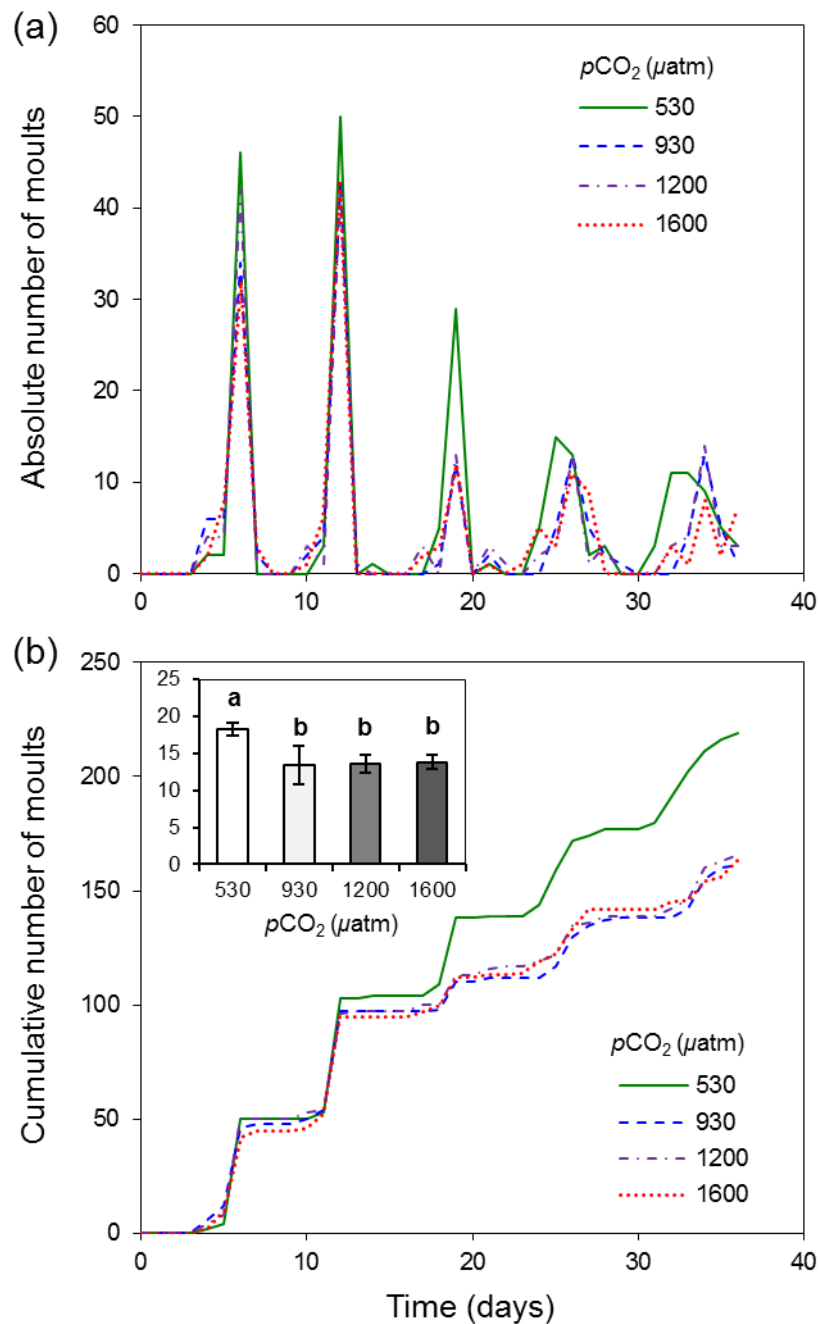


Fig 3 (a) Absolute moult number and (b) cumulative moult number of all juvenile *P. flexuosus* individuals over time (low and high food combined, see Fig. S4 for graphs of both low and high food treatments). Inlet in (b) shows the mean \pm SD ($n=3$) cumulative number of moults per $p\text{CO}_2$ level at the end of the experiment (different letters indicate significant differences among $p\text{CO}_2$ levels, Tukey contrasts, $P < 0.05$).

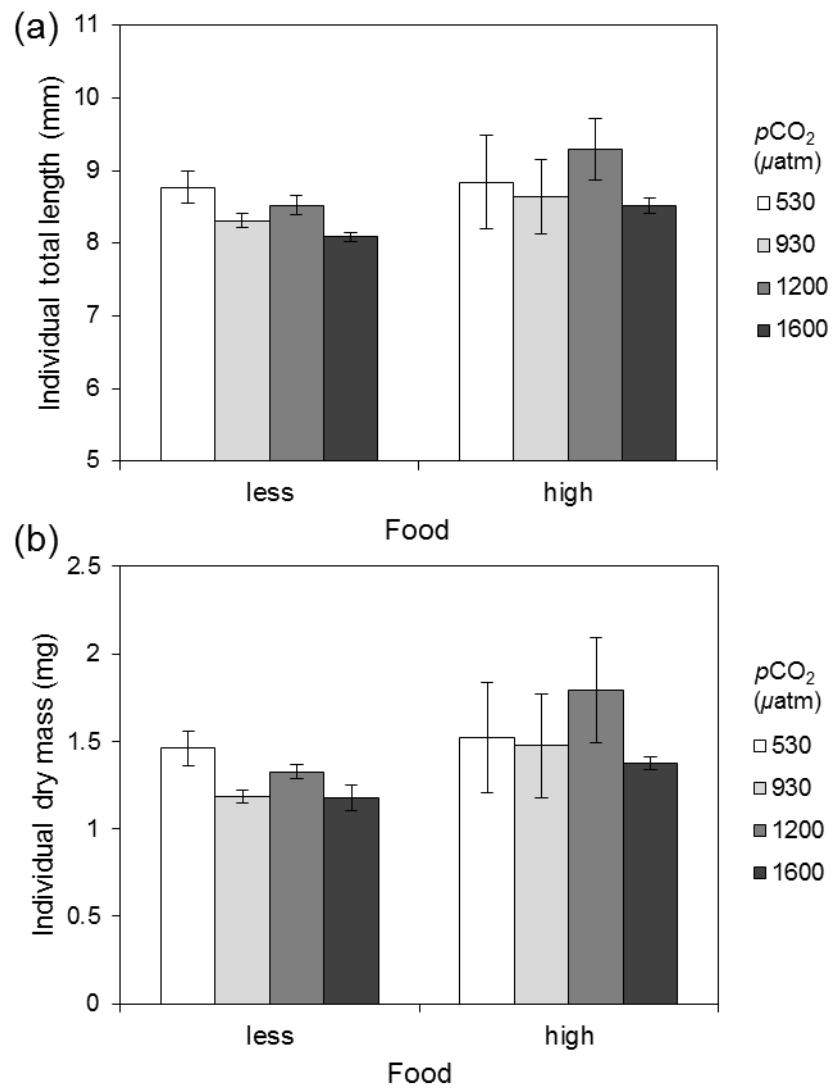


Fig 4 Mean \pm SD (n=3) for (a) total length (mm) and (b) dry mass (mg) of juvenile *P. flexuosus* after exposure of 5 weeks to different $p\text{CO}_2$ levels in the treatments of low and high food.

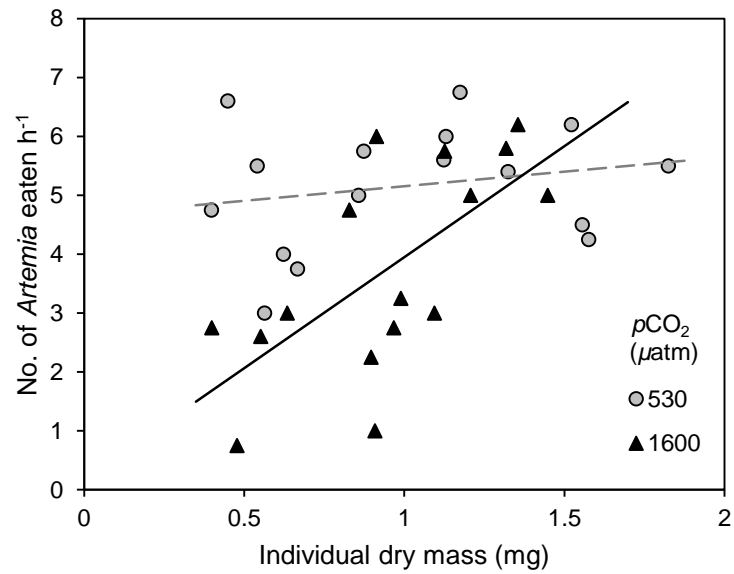


Fig 5 Short-term feeding rates of juvenile *P. flexuosus* after exposure of 5 weeks to 530 μatm (ambient) or 1600 μatm (high) $p\text{CO}_2$ depending on dry mass of individual mysids (see Fig. S5 for feeding rates depending on individual total length). Feeding rates were measured as number of *Artemia salina* nauplii eaten per hour and determined on 3 consecutive days at the end of the juvenile experiment using 5-6 individuals per day from additional juveniles kept at high food supply.

Table 1. Carbonate chemistry of experimental seawater in exposure tanks.

pH transmitter setting	8.0 (amb.)	7.75	7.6	7.5
Temperature (°C)	12.4 ± 0.1	12.5 ± 0.2	12.6 ± 0.2	12.6 ± 0.2
pH (total scale)	7.94 ± 0.01	7.72 ± 0.01	7.61 ± 0.01	7.50 ± 0.02
<i>p</i> CO ₂ (μatm)	525.0 ± 14.1	925.6 ± 27.6	1208.8 ± 28.8	1588.6 ± 83.4
C _T (μmol kg ⁻¹)	2160.4 ± 4.0	2242.0 ± 3.8	2276.5 ± 3.6	2311.1 ± 7.1
HCO ₃ ⁻ (μmol kg ⁻¹)	2019.0 ± 5.8	2128.6 ± 4.5	2167.3 ± 3.8	2199.6 ± 6.1
CO ₃ ²⁻ (μmol kg ⁻¹)	120.2 ± 2.3	76.1 ± 1.8	60.5 ± 1.5	47.6 ± 2.5
CO ₂ (μmol kg ⁻¹)	21.2 ± 0.5	37.3 ± 1.1	48.7 ± 1.3	63.9 ± 3.5
ΩCa	2.86 ± 0.06	1.81 ± 0.04	1.44 ± 0.04	1.13 ± 0.06
ΩAr	1.83 ± 0.04	1.16 ± 0.03	0.92 ± 0.02	0.72 ± 0.04

Average values ± SD (n = 17) calculated using CO₂sys (see Materials and Methods for details) with measured temperature, pH, total alkalinity (A_T, 2319.2 ± 8.3 μmol kg⁻¹, n = 6), salinity (35.04 ± 0.06‰, n = 17), phosphate (1.39 ± 0.64 μmol kg⁻¹, n = 6), and silicate (5.89 ± 0.08 μmol kg⁻¹, n = 6) as input variables. Approximate *p*CO₂ values (highlighted in bold) were used to indicate *p*CO₂ levels of this study (i.e. 530, 930, 1200, and 1600 μatm).

Table 2. Statistical results of mixed effects Cox models with tank-id as a random effect and food regime and $p\text{CO}_2$ level as fixed effects (individuals nested within jar).

	χ^2	df	<i>P</i>
juvenile mysids			
high and low food			
$p\text{CO}_2$	2.88	3	0.411
food	0.032	1	0.857
$p\text{CO}_2 \times \text{food}$	5.82	3	0.121
low food			
$p\text{CO}_2$	1.89	3	0.596
high food			
$p\text{CO}_2$	7.58	3	0.056

Table 3. ANOVA type III results for fixed effects of linear mixed models fitted to the cumulative number of moults at the end of the experiment, inter-moult period between the fourth and fifth moult, mean total length (mm), and mean dry mass (μg) with $p\text{CO}_2$ (μatm) and food regime as fixed effects (and tank-id as a random effect).

	df (num,denum)	<i>F</i>	<i>P</i>	
Cumulative number of moults				
$p\text{CO}_2$	3,48	7.42	0.0004	***
food	1,48	0.02	0.88	
$p\text{CO}_2 \times \text{food}$	3,48	0.68	0.57	
Inter-moult period 4 th -5 th moult				
$p\text{CO}_2$ (cont.)	1,7	10.6	0.014	*
food	1,25.9	0.90	0.351	
$p\text{CO}_2 \times \text{food}$	1,26.5	2.63	0.117	
Total length				
$p\text{CO}_2$	3,48	4.54	0.007	**
food	1,48	9.92	0.003	**
$p\text{CO}_2 \times \text{food}$	3,48	1.43	0.247	
$\log_{10}(\text{dry mass})$				
$p\text{CO}_2$	3,48	3.80	0.016	*
food	1,48	16.01	0.0002	***
$p\text{CO}_2 \times \text{food}$	3,48	1.59	0.203	

Note that $p\text{CO}_2$ was set to a continuous fixed effect in the model fit with inter-moult period and not factorial as in the other model fits. Degrees of freedom (df) are calculated using Satterthwaite approximations and thus can be non-integers. Significant effects are indicated by asterisks: * $P < 0.05$, ** $P < 0.01$ and *** $P < 0.001$.

Table 4. Intermoult periods (in days) observed in the experiment using juvenile *P. flexuosus* exposed to different $p\text{CO}_2$ (μatm) levels.

$p\text{CO}_2$ (μatm)	to 1st	1st-2 nd	2nd-3rd	3rd-4th	4th-5th
low food					
530	6.96 ± 0.07	6.00 ± 0.00	6.95 ± 0.20	6.45 ± 0.25	7.59 ± 0.22*
930	5.68 ± 0.16	6.10 ± 0.07	7.14 ± 0.25	7.14 ± 0.75	7.92 ± 0.12*
1200	5.77 ± 0.23	6.28 ± 0.20	7.13 ± 0.22	7.67 ± 1.33	8.54 ± 0.65*
1600	5.75 ± 0.54	6.16 ± 0.39	6.92 ± 0.15	7.18 ± 0.16	8.77 ± 0.88*
high food					
530	5.84 ± 0.15	6.16 ± 0.15	6.86 ± 0.43	6.93 ± 1.21	7.64 ± 0.46
930	5.66 ± 0.09	6.24 ± 0.19	7.44 ± 0.51	6.85 ± 0.60	7.60 ± 0.15
1200	5.70 ± 0.26	6.07 ± 0.21	7.07 ± 1.78	6.25 ± 1.16	7.92 ± 0.12
1600	5.74 ± 0.13	6.03 ± 0.04	6.23 ± 0.04	7.02 ± 1.44	7.89 ± 0.79
low and high food combined					
530	5.90 ± 0.10	6.08 ± 0.07	6.91 ± 0.23	6.69 ± 0.66	7.61 ± 0.21*
930	5.67 ± 0.12	6.17 ± 0.12	7.35 ± 0.33	6.92 ± 0.23	7.76 ± 0.01*
1200	5.73 ± 0.23	6.17 ± 0.05	7.12 ± 0.72	6.96 ± 1.21	8.14 ± 0.09*
1600	5.75 ± 0.29	6.09 ± 0.19	6.72 ± 0.29	7.25 ± 0.56	8.39 ± 0.64*

Average values ± SD (n=3) are given; numbers in parentheses indicate number of observations/jars. * indicates a significant increase along the $p\text{CO}_2$ gradient ($p\text{CO}_2$ as continuous variable low food: $F(1,19) = 11.15$, $P = 0.0035$; low and high food combined: $F(1,41) = 10.89$, $P = 0.0020$).