## REPORT OF THE

# WORKING GROUP ON NORTH ATLANTIC SALMON <br> Part one 

Aberdeen, Scotland
2-11 April 2001

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### 1.1 Main Tasks

At its 2000 Statutory Meeting, ICES resolved (C. Res. 2000/2ACFM07) that the Working Group on North Atlantic Salmon [WGNAS](Chair: Dr. N. Ó Maoiléidigh, Ireland) will meet in Edinburgh, UK from the 2-11 April 2001 to consider questions posed to ICES by the North Atlantic Salmon Conservation Organisation (NASCO). An alternative venue was selected in Aberdeen, UK during the same period. The terms of reference and sections of the report in which the answers are provided, follow.

| a) With respect to Atlantic salmon in the North Atlantic area: | Section |
| :---: | :---: |
| i. provide an overview of salmon catches and landings, including unreported catches by country and catch and release, and worldwide production of farmed and ranched salmon in 2000; | 2.1 \& 2.2 |
| ii. report on significant developments which might assist NASCO with the management of salmon stocks; | 2.4, 3.9 |
| iii. use case studies to illustrate options for taking account of risk in the provision of catch advice and comment on the relative merits of each option; | 2.3 |
| iv. assess the possible reasons for the differences in the occurrence of escaped farmed fish in fisheries and stocks in different areas; | 2.4 |
| v. advise on the potential biases in the catch advice model resulting from the inclusion of fish farm escapes in the assessment models; | 3.6 |
| vi. provide a compilation of tag releases by country in 2000. | 2.5 |
| b) With respect to Atlantic salmon in the North-East Atlantic Commission area: |  |
| i. describe the events of the 2000 fisheries and the status of the stocks; | 3.1-3.4 |
| ii. update the evaluation of the effects on stocks and homewater fisheries of significant management measures introduced since 1991; | 3.5 |
| iii. further develop the age-specific stock conservation limits where possible based upon individual river-based stocks; | 3.7 |
| iv. provide catch options or alternative management advice with an assessment of risks relative to the objective of exceeding stock conservation limits; | 3.8 |
| v. update the information on by-catch of salmon post-smolts in pelagic fisheries; | 3.9 |
| vi. identify relevant data deficiencies, monitoring needs and research requirements. | 3.10 |
| c) With respect to Atlantic salmon in the North American Commission area: | Section |
| i. describe the events of the 2000 fisheries and the status of the stocks; | 4.1 \& 4.2 |
| ii. update the evaluation of the effects on US and Canadian stocks and fisheries of management measures implemented after 1991 in the Canadian commercial salmon fisheries; | 4.3 |
| iii update age-specific stock conservation limits based on new information as available; | 4.4 |
| v. provide catch options or alternative management advice with an assessment of risks relative to the objective of exceeding stock conservation limits; | 4.5 |
| v. identify relevant data deficiencies, monitoring needs and research requirements. | 4.6 |


| d) With respect to Atlantic salmon in the West Greenland Commission area: | Section |
| :--- | :--- |
| i. describe the events of the 2000 fisheries and the status of the stocks; | $5.1 \& 5.2$ |
| ii. update the evaluation of the effects on European and North American stocks of the Greenlandic <br> quota management measures and compensation arrangements since 1993; | 5.4 |
| iii. provide a detailed explanation and critical examination of any changes to the model used to <br> provide catch advice and of the impacts of any changes to the model on the calculated quota; | 5.5 |
| iv. provide catch options or alternative management advice with an assessment of risks relative to <br> the objective of exceeding stock conservation limits; | 5.7 |
| v. evaluate potential causes for the changes in the Continent of origin of salmon captured in the <br> West Greenland fishery including potential changes in marine migration patterns; | 5.3 |
| vii. identify relevant data deficiencies, monitoring needs and research requirements. | 5.9 |

The Working Group considered 31 Working Documents submitted by participants (Appendix 1); other references cited in the report are given in Appendix 2.

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A full address list for the participants is provided in Appendix 3.

### 2.1 Catches of North Atlantic Salmon

### 2.1.1 Nominal catches of salmon

The nominal catch of a fishery is defined as the round, fresh weight of fish which are caught and retained. Total nominal catches of salmon reported by country in all fisheries for 1960-2000 are given in Table 2.1.1.1. Catch statistics in the North Atlantic also include fish farm escapees and, in some north-east Atlantic countries, ranched fish (see Section 3).

The Icelandic catches are presented under two separate categories; wild and ranched. Iceland is the only North Atlantic country where large scale ranching has previously been undertaken and where the intent was to harvest all returns at the release site. While ranching does occur in other countries it is on a much smaller scale. Some of these operations are experimental and at others harvesting does not occur solely at the release site. The ranched component in these countries has therefore been included within a single figure for the nominal catch.

Figure 2.1.1.1 shows the nominal catch data grouped by the following areas: 'Scandinavia and Russia' (including Denmark, Finland, Iceland, Norway, Russia and Sweden); ‘Southern Europe’ (including Spain, France, Ireland, UK (England and Wales), UK (Northern Ireland) and UK (Scotland)); and 'North America' (including Canada, USA and St Pierre et Miquelon); and 'Greenland and Faroes'.

The provisional total nominal catch for 2000 is 2814 t , which is the highest since 1996 . This catch is 568 t greater than the updated catch for 1999 (2246t) and although greater than the previous 5 -year average (2754t) , is 636t less than the previous 10 -year average ( 3450 t ). In all, 10 countries reported an increase in the 2000 catch compared to the final 1999 values. Catches in 10 countries were greater than the previous 5 -year averages and catches in 5 were greater than previous 10-year averages.

Several countries partition reported nominal catches by size or sea-age category and these data, where available, are given in Tables 2.1.1.2 and 2.1.1.3. The figures for 2000 are provisional and, as in Table 2.1.1.1, catches in some countries include both wild and reared salmon (excluding ranched fish from Iceland) and fish farm escapees. Different countries use different methods to partition their catches by sea-age class and these methods are described in the footnotes to Table 2.1.1.3. The composition of catches in different areas is discussed in more detail in Sections 3, 4 and 5.

Table 2.1.1.4 presents, where data are available, the nominal catch by country partitioned according to whether the catch was taken by coastal, estuarine or riverine fisheries. In addition, fisheries in West Greenland, Faroes and St. Pierre et Miquelon are exclusively coastal or on the high seas. The proportions accounted for by each fishery varied considerably among countries although overall proportions remained relatively stable. In total, coastal fisheries accounted for $53 \%$ of catches in North East Atlantic countries in 2000 compared to $52 \%$ in 1999, whereas in-river fisheries took $41 \%$ of catches in both 1999 and 2000. In North America, coastal fisheries accounted for $9 \%$ of the catch in 2000 compared to $7 \%$ in 1999, while in-river fisheries took $77 \%$ of catches in 2000 compared to $67 \%$ in 1999.

### 2.1.2 Catch and release

The practice of catch and release (often termed hook and release) in rod (recreational) fisheries has been used as a conservation measure for salmon in some areas of Canada and USA since 1984. Recent declines in salmon abundance in the North Atlantic have resulted in an increased use of this management option., either as a voluntary practice or through statutory regulation. The nominal catches presented in Section 2.1.1 are comprised of fish which have been caught and retained and do not include catch-and-release salmon. Table 2.1.2.1 presents catch-and-release information from 1991-2000 for those countries that have records. Catch-and-release may be practiced in other countries while not being formally recorded. There are large differences in the percentage of the total rod catch that is released reflecting the varying management practices among these countries. Within countries, however, this percentage has tended to increase in recent years. Thus in 2000, although release rates range from approximately $10 \%$ in Iceland to $74 \%$ in Russia, rates in 2000 are among the highest in each 10-year series for most countries.

### 2.1.3 Unreported catches

Unreported catches by year and Commission Area are presented in Table 2.1.3.1. A description of the methods used to evaluate the unreported catches was provided in ICES 2000/ACFM:13. The 2000 unreported catch can be compared to
previous years values as the estimation method used by each country is relatively unchanged. However, it may not be appropriate to compare the unreported catch of one country to another as the same information may not be included in the estimate. For example, some countries include only the illegal landings in the unreported catch, while other countries include unreported legal catch and illegal catches in their estimates and the illegal catch is included with the nominal catch for France.

The total unreported catch in NASCO areas in 2000 was estimated to be $1,269 \mathrm{t}$, an increase of $23 \%$ from the 1999 estimate. Estimates were derived for the North American Commission Area (124t), the West Greenland Commission Area ( 10 t ) and North East Atlantic Commission Area ( $1,135 \mathrm{t}$ ). Figure 2.1.3.1 shows that the unreported catch has remained a relatively constant proportion (30\%) of the total catch since 1987. No data for the combined three Commission Areas are available prior to 1987. Where available, data are presented by country for 2000 (Table 2.1.3.2). The individual inputs to the total North Atlantic catch range from $0 \%$ to $16 \%$. While this broadly indicates the level of unreporting by each country relative to the total catch in the North Atlantic, it should be noted that these estimates are not precise and are difficult to validate. The percentage of the total national catches (reported + unreported) by country ranges from $0 \%$ to $67 \%$.

It is not known whether any vessels fished for salmon in the international waters in the Norwegian Sea. There were no surveillance flights reported to have been undertaken by the Icelandic and Norwegian Coastguards over the winter period 2000/2001 when fishing for salmon would be most likely to occur.

### 2.2 Farming and Sea Ranching of Atlantic Salmon

### 2.2.1 Production of farmed Atlantic salmon

The production of farmed Atlantic salmon in the North Atlantic area was $658,735 \mathrm{t}$, in 2000 (Table 2.2.1.1 and Figure 2.2.1.1), an increase in production over 1999 ( $636,783 \mathrm{t}$ ). The 2000 production was $30 \%$ higher than the 1995-99 average ( $504,809 \mathrm{t}$ ) for the area. The countries with the largest production were Norway and Scotland, accounting for $65 \%$ and $20 \%$ of the reported North Atlantic total. Reported increases compared to average production for 1995 to 1999 (Table 2.2.1.1) ranged from $75 \%$ for eastern Canada to $6 \%$ for UK(N. Ireland).

The worldwide production of farmed Atlantic salmon in 2000 was $704,134 \mathrm{t}$ (excluding Chile; Table 2.2.1.1 and Figure 2.2.1.1). Outside the North Atlantic area, data were only compiled for production of farmed Atlantic salmon in western North America (Washington and British Columbia), where 2000 production was greater than 1999. The worldwide production of farmed Atlantic salmon compiled for 2000 was over 200 times the reported nominal catch of Atlantic salmon in the North Atlantic. As a result, aquaculture fish dominate world markets, and have probably contributed to the decline in commercial fishing effort in many countries.

### 2.2.2 Production of ranched Atlantic salmon

Ranching has been defined as the production of salmon through smolt releases with the intent of harvesting the total population that returns to freshwater (harvesting may include collecting fish for broodstock) (ICES 1994/Assess:16). The total production of ranched Atlantic salmon in countries bordering the North Atlantic in 2000 was $11 \mathrm{t}, 22 \mathrm{t}$ lower than in 1999 ( 33 t ) and the lowest value since 1984 (Table 2.2.2.1 and Figure 2.2.2.1). Production in Iceland declined dramatically because no smolts were released into ocean ranching in 1999, thus, only 2SW fish were harvested in 2000. Production of ranched fish was less than 5 t in each of the three other countries reporting (Ireland, UK(N. Ireland), and Norway). Production in these three countries includes catches in net, trap and rod fisheries. Icelandic catches, on the other hand, are entirely from estuarine and freshwater traps at the ranching stations.

### 2.3 Use of Case Studies to Illustrate Options for Taking Account of Risk in the Provision of Catch Advice

The Working Group considered this question, together with the supplementary request that "ICES provide information that will assist with the implementation and evaluation by NASCO and its Contracting parties of the decision structure (Annex 4 of document $\mathrm{CNL}(00) 18$, provisionally adopted by the Council).

Management of Atlantic salmon in the North American and Greenland Commission areas is based on a fixed escapement strategy. All potential recruits in excess of the conservation requirement are considered to be available for harvest. The conservation requirement is considered to be a threshold reference point. The undesirable event is that the spawning escapement to North America will be below the conservation limit. The probability of achieving the biological objective (conservation requirement) depends on the quota selected, the uncertainties of the harvest, and the uncertainties of the forecast. The level of quota selected by managers should be based on the level of risk that they
consider acceptable. The greater uncertainty in the forecast and harvest the lower the target exploitation or quota would have to be to attain the same risk acceptance level. The exploitation rate, the quota available at this risk acceptance level or the quota plus the spawner escapement reserve could be considered as management targets for that fishery for that year.

The current management approach used or the West Greenland fishery considers the catch options relative to a $50 \%$ probability of achieving the conservation limit (or a $50 \%$ chance of the undesirable event occurring) and ignores the uncertainty in the stock assessment. Ignoring uncertainty is inconsistent with the principles defined under the precautionary approach. The provision of catch advice in a risk framework involves the incorporation of the uncertainty in all the factors contributing to the assessment of stock status, development of the forecast, and fisheries management.

Risk analysis is useful for making decisions in uncertain situations and may be used to select management options that convey lower levels of risk. Risk analysis for Atlantic salmon is conducted with respect to achieving a conservation limit or combinations of conservation limits. Analysis is based on uncertain exploitation rates, harvest components and imprecise forecasts.

Risk management for mixed stock fisheries is complex. A prime consideration for achieving the biological objective in all stock components is the minimum acceptable probability of achieving the objective for the smallest component. Based on acceptance of this probability, a target exploitation rate or quota can be set from the uncertainty distributions of the forecast and exploitation.

The Working Group considered case studies to illustrate two approaches for taking account of risk in the provision of catch advice. The first considers incorporating the uncertainties in a risk analysis to provide a probability profile of meeting the conservation objective. The second approach addresses the use of management targets to increase the probability of meeting the conservation objective.

### 2.3.1 Case studies for calculating risk for the provision of catch advice.

The case studies below consider a mixed stock fishery example for two levels of abundance and a single stock example for a homewater fishery.

## Case Study 1 Mixed stock fishery - West Greenland fishery

The deterministic calculation of catch options for the West Greenland fishery uses the point estimates of the input parameters. When the input parameters have uncertainty (for example the PFA value), the value at the $50 \%$ probability level is used. The procedure is described in Appendix 6 and summarized for two stock levels (low and moderate) (Figure 2.3.1.1). For the low abundance period, the catch option at the point estimates and for a $40 \%$ allocation of surplus to West Greenland was 19 tons. For the period of higher abundance, the deterministic catch option was 561 t .

The deterministic calculation does not provide any analysis of the risk of achieving conservation requirements in North America at the calculated catch level. The data inputs are uncertain and a risk analysis for the objective of achieving the conservation limits must appropriately incorporate these uncertainties. The uncertainties included in these case studies are:

- Conservation requirement uncertainty for six stock areas
- Uncertainty in the forecast PFA value
- Uncertainty in the biological characteristics of the salmon in the fishery

Management error (for example, not catching the exact quota) has not been incorporated but it could be included if an estimate was made from historic data.

### 2.3.1.1 Conservation requirement uncertainty.

The 2SW spawner requirement for North America used by ICES is the sum of point estimates of individual river or fishing area spawner requirements. It has been shown that the sum of individual river or area requirements provides a probability level of less than $50 \%$ of simultaneously meeting the conservation requirements in individual rivers or areas (ICES 1996/Assess:11). This excludes the uncertainty in the individual river reference points which have not been quantified for the majority of rivers of North America (except for Quebec rivers). The sum of the 2 SW spawner requirements for North America is 152,548 fish and adjusted for natural mortality to the point prior to the fishery
(spawner reserve) is 170,286 fish. To ensure a spawner escapement of $100 \%$ of conservation into six stock areas simultaneously at a $50 \%$ probability level, $169,0002 \mathrm{SW}$ are required to return to North America, 188,650 2SW fish released from the fishery (Figure 2.3.1.2). The uncertainty increases as the number of stock areas defined by managers increases. Additionally, the analysis assumes that the stock areas are all producing at the same rate relative to their conservation limits.

| Probability of achieving conservation requirements simultaneously in six stock areas |  |  |
| :--- | :--- | :--- |
|  |  | Prior to fishery (adjusted by M |
| Probability Level | In North America | for 11 months) |
| Point estimate | 152,548 | 170,286 |
| $50 \%$ | 169,000 | 188,650 |
| $75 \%$ | 173,000 | 193,100 |
| $90 \%$ | 177,000 | 197,600 |

### 2.3.1.2 PFA Forecast Uncertainty

Forecasts of abundance in the year to come are dependent upon stochastic functional relationships. Generally, the forecasts have large uncertainty. The PFA forecasts for the low abundance and moderate abundance periods are shown in Figure 2.3.1.3. The PFA forecast value at a $50 \%$ level for the moderate abundance period was 437,000 fish compared to 183,000 fish in the low abundance period.

|  | Forecast values |  |
| :--- | :--- | :--- |
| Probability level | Moderate abundance | Low abundance |
| $10 \%$ | 236,782 | 120,000 |
| $30 \%$ | 342,213 | 155,000 |
| $50 \%$ | 436,770 | 183,000 |
| $70 \%$ | 553,223 | 215,000 |
| $90 \%$ | 801,849 | 280,000 |

### 2.3.1.3 Biological characteristics of the fish

Biological characteristics of the fish in the fishery of the coming year are also unknown. These are estimated based on characteristics of previous years taking account of any temporal trends in characteristics if they occur. In the deterministic approach, the point estimates (such as the average weight of previous years, the average fecundity of females in recent years) are used. In the risk analysis, the uncertainty in the characteristics are considered. The four characteristics are: proportion North American origin 1SW salmon, mean weight of 1SW salmon of North America, the mean weight of 1SW salmon of European origin, and the age correction factor for older age groups in the fishery. The variability in the number of 1 SW salmon at a given catch option is illustrated in Figure 2.3.1.4. For a catch of 50 tons, the expected catch of 1 SW salmon of North American origin can vary between 11,700 fish and 15,450 fish $\left(10^{\text {th }}\right.$ to $90^{\text {th }}$ percentiles).

### 2.3.1.4 Completing the Risk Analysis

Incorporating all these uncertainties results in a measure of the reliability of the stock assessment for making management decisions. The reliability of the assessment has different and profound consequences on the catch options considerations.

In the theoretical example shown in Figure 2.3.1.5, two assessments provide the same point estimate ( $50 \%$ probability value) but the precisions are very different. Under a risk-prone management approach, the allowed catch would be greater for the imprecise assessment: at a $70 \%$ risk level, the advised catch under the precise assessment would be 500 t but the uncertain assessment would provide for a catch of 800 t . The risk-averse management approach would advise for lower catch options for the imprecise assessment: at a $20 \%$ risk level, the precise assessment would provide a catch option of about $400 t$ but for the imprecise assessment, no catch is advised. A risk-averse approach is a pre-requisite of precautionary management.

The risk analysis probability profiles for the two years of contrasting stock abundance are shown in Figure 2.3.1.6. In the higher abundance year, a catch option of about 1,250 tons produces a $50 \%$ probability level of achieving the conservation requirement. This contrasts with the low abundance year when a catch option of just over 100 tons
provides a $50 \%$ probability of meeting conservation. Note that these catch levels are higher than the deterministic calculations because high prefishery abundance values cannot be discounted. To adopt a more risk-averse approach, managers must select a higher probability (or lower risk of stocks) exceeding their conservation limits. At a $65 \%$ probability level, there would not be any available harvest in the low abundance scenario and a quota of about 900 t in the higher abundance scenario.

The risk analysis described above has not incorporated management uncertainty. When management is imperfect, as is generally the case, the effect on the risk analysis is to increase the uncertainty in the probability of meeting the conservation objectives. The analysis has also excluded any differences in status among the stock areas. In the case where stock status differs, the probability of meeting conservation for a given year will be overestimated since the spawning escapement to the areas will be different from those assumed in the model. An evaluation of performance of previous year fisheries would provide valuable insight into the appropriateness of the data inputs and the assumptions of the risk analysis.

## Case Study 2 Single River Example - Miramichi River

The Miramichi River, at a maximum axial length of 250 km and draining an area of about $14,000 \mathrm{~km}^{2}$, has the largest Atlantic salmon run of eastern North America. There are two major branches: the Northwest Branch covers about 3,900 $\mathrm{km}^{2}$ and the Southwest Branch about $7,700 \mathrm{~km}^{2}$ of drainage area. The two branches drain into a common estuary and subsequently drain into the Gulf of St. Lawrence at latitude $47^{\circ} \mathrm{N}$. Separate branch assessments were introduced in 1992 to account for the differences in exploitation between the Northwest and Southwest branches. Native Peoples fisheries were historically conducted almost exclusively in the Northwest Miramichi (exploitation also occurs in the estuarial waters of the Miramichi River, downstream of the confluence of the two branches) and recreational fisheries exploitation also differs between the Northwest and Southwest branches.

Temporal stock distinctiveness has also been highlighted as an important component of the Atlantic salmon resource of the Miramichi. Early runs and late runs have different composition in terms of small and large salmon proportions and sex ratios. The early runs in both branches are also exploited more heavily than the late runs.

Atlantic salmon are presently exploited in Native Peoples and recreational fisheries. No large salmon ( $>=63 \mathrm{~cm}$ fork length) can be retained in the recreational fishery (mandatory catch and release) and Native Peoples fisheries for small and large salmon are under gear, season and quota controls.

The conservation spawning requirement for the Miramichi River and each branch separately is based on an egg requirement of $2.4 \mathrm{eggs} / \mathrm{m}^{2}$ of spawning and rearing habitat area. The objective is to obtain all the egg depositions from large salmon although compliance relative to the achievement of the conservation requirement is determined relative to egg depositions from both small and large salmon.

### 2.3.1.5 Forecast of returns in 2001

The association between small salmon (almost exclusively 1SW salmon) and large salmon returns the subsequent year was examined over the time series, 1985 to 2000 (Figure 2.3.1.7). The ratio of small salmon to large salmon for this time period varied between 1.4 and 7.1 with the most recent year ratio ( 1999 small, 2000 large salmon) at 1.41 . The median ratio model for the recent five-year period (1995 to 1999) would predict returns of large salmon (including previous spawners) of 16,400 fish (ranging between 14,700 and 25,200).

### 2.3.1.6 Risk analysis of the fishery

The probability of meeting conservation requirements in 2001 was estimated from the predicted return of large salmon in 2001 based on the small:large salmon ratio of 1996 to 2000 and assuming that small salmon returns in 2001 would be similar to the previous five-year average. The model to assess the risk to conservation if fisheries were to occur in year 2001 can account for seasonal differences in harvest levels, catch-and-release mortality, and biological characteristics of the adults (Figure 2.3.1.8).

Risk is quantified in terms of the probability of meeting conservation and the egg loss resulting from the fisheries harvests as a percentage of total eggs in the returns of adult salmon to the river. For the Miramichi River overall, there is a $54 \%$ probability of meeting conservation in year 2001, in the absence of fisheries (Figure 2.3.1.9). Egg loss as a percentage of total eggs in the returns would be less than $5 \%$ and the probability of meeting the conservation objectives would be $45 \%$ if large salmon losses due to fisheries (regardless of user) were less than 1000 fish and small salmon losses were less than 9000 fish (Figure 2.3.1.9).

### 2.3.2 Case studies for use of management targets as a means of minimizing risk

NASCO (1998) proposed that "stocks be maintained above conservation limits by means of management targets". The purpose of the management target would be to satisfy the management objective of ensuring a high probability that the conservation limit will be exceeded. Targets are aim points. In 2000, the Working Group acknowledged that it was the responsibility of managers to define the level of risk, resulting from uncertainty, of stocks falling below the conservation limit (ICES CM 2000/ACFM: 13). Once the level of risk is defined, it may be possible to set a management target.

Within the case study examples provided in the draft decision structure, the use of management targets at some value proportionally higher than the conservation limit was used. The challenge is to assess whether a management target can be defined which would provide a consistent increase to the probability of meeting the conservation requirement.

The derivation of a mangement target to increase the probability of meeting the conservation requirements can be based upon an assessment of the same uncertainties as the previous approach, namely conservation requirement uncertainty for six stock areas, uncertainty in the forecast PFA value, and uncertainty in the biological characteristics of the salmon in the fishery.

The use of a management target assumes that managers may choose to harvest all the surplus, and the risks are therefore assessed on the assumption that this will be done. The analysis provided in Section 2.3.1 illustrates how a management target could be set which would increase the probability of achieving the stock conservation limits by a fixed amount. For the low abundance year, a probability of achieving the conservation requirement of $60 \%$ would result from a management target set at $116 \%$ of the conservation requirement which would result in a foregone harvest of 100 t (Figure 2.3.1.6). For the higher abundance year, a management target of $128 \%$ of the conservation requirement would be required to achieve a probability of meeting conservation of $60 \%$ and this target would result in a foregone harvest of 250 t (from 1,250 t at $50 \%$ to $1,000 \mathrm{t}$ at $60 \%$ ) (Figure 2.3.1.6).

The problem with this approach is that the same management target (a fixed proportion above the conservation limit) will not have a consistent effect in reducing risk. Although an average value could be employed, it could result in significant over-exploitation (or foregone harvest) in different years.

### 2.3.3 Relative merits of the approaches

Our analysis shows that there is no single management target level which will provide the same level of risk of failing to meet the conservation objectives over variations in abundance and assessment uncertainty. The Working Group therefore favours the approach of providing an annual risk analysis which considers the variations in abundance, and in the uncertainty of the assessment. The deterministic calculations for any year can be completed by the managers and the resultant probability of meeting the conservation objective at the calculated catch level is visually derived from the risk analysis probability profile plot (such as in Figure 2.3.1.6). The Working Group recognizes that it is the responsibility of managers to select the level of risk of stocks falling below the conservation limit and emphasizes that the appropriate risk level might be different for different fisheries; mixed stock fisheries pose a greater risk to conservation of individual salmon populations than single river stock fisheries.

### 2.3.4 Review of draft decision structure (NASCO CNL(00)18)

The Working Group tabled the Report of the Standing Committee on the Precautionary Approach- (CNL(00) 18 Application of a Precautionary Approach to Management of Salmon Fisheries) for discussion and comment.

The Group considered that the draft decision structure provided a very useful first step in developing mechanisms for guiding managers towards appropriate actions for fisheries, compatible with the underlying goal that conservation requirements (both abundance and diversity) of contributing stocks is achieved. The Working Group endorsed the emphasis given in the draft decision structure to systematically monitoring the effect of management measures and taking results into account in future management decisions. It was also felt that the various elements of the decision structure, if widely applied to fisheries and stocks, would provide a useful audit trail, showing the data available for stocks and the basis of the management decisions taken for the fisheries where those stocks are represented. This would also provide clear indications of data deficiencies and highlight where lack of data was impeding sound management.

The Group was concerned by the absence of any clear indication of how the structure was meant to be used, as both questions and instructions were included. The presentation would be improved by adopting a $\mathbf{F}_{\text {low }}$ diagram type of approach, similar to that provided in the working group report (ICES CM 2000/ACFM: 13). This should make the review and evaluation of measures taken more explicit (by means of feedback loops) and should indicate where where risk should be considered.

The step in the single stock framework that refers to stocks threatened by external factors is unclear, as it is not obvious what happens if the stock is threatened but is not yet below the conservation limit (for example recently introduced disease into a still productive stock). It may be better to incorporate this into the general assessment of status, such that if status is threatened by external factors, the reasons could be identified and appropriate pre-agreed management actions taken.

The Group noted the clear distinction between action under conditions of unsatisfactory stock status (i.e. identify reasons and implement corrective action) and actions under conditions of surplus (implement pre-agreed management actions to harvest the surplus). However, the Group felt it was likely that many stocks with an exploitable surplus are also subject to impacts that may cause them to fall below surplus at some future time, if measures to mitigate impacts (for example, habitat rehabilitation) are not implemented. Therefore, it was insufficient to recommend implementation of measures only when status had become fully unsatisfactory.

The Working Group noted that pre-agreed management actions should take account of all sources of uncertainty, with management targets being suggested where appropriate, however the draft decision framework did not fully address the incorporation of risk into the decision process. In this respect, the use of further case studies specifically to illustrate this would be valuable (see Section 2.3.1).

It was noted that no pre-agreed management actions were specified, though it is accepted that a generic structure may not be able to cover all specific cases.

In summary, the Working Group recommends some modifications and reference to similar salmon management structures being developed by contracting parties for use in homewater fisheries.

### 2.4 Significant development towards the management of salmon

### 2.4.1 Infectious salmon anaemia: implications for wild salmon management

Information was presented to the Working Group about infectious salmon anaemia (ISA) in North America.

ISA has caused extensive mortalities at salmon farms. The disease has been reported from the industry in Norway (1984), East Coast Canada (1996), Scotland (1999) and the Faroes (2000). Positive tests in wild fish were obtained in Canada and Scotland in (1999). In 2000 the Working Group expressed concern about the spread of the disease within wild populations, and the subsequent mortalities that could result in wild fish.

In Canada, aggressive control measures taken by the East Coast salmon farming industry seem to be working. At present, only one site has reported the disease in the smolt class that was transferred to the sea cages in spring 2000. No ISA was detected in wild and escaped-farmed fish entering the Magaguadavic River in 2000, where positive tests for both groups were obtained for the first time in 1999.

Initial reports in 2000 of the presence of the virus for the first time in the Margaree River in Nova Scotia ( 2 of 30 fish tested), the Morell River in Prince Edward Island (4 of 30 fish), and the Saint John River New Brunswick (16 of 36 fish) are problematic because they could not be confirmed with additional testing. The Working Group remains concerned about the potential spread of this disease.

The first confirmed case of ISA from the East Coast USA salmon farming industry was announced on 16 March 2001. The US industry is now implementing measures similar to those used in Norway, Scotland, and Canada to manage the problem.

Recent genome comparisons of European and Canadian strains of the virus found Scottish and Norwegian strains were $98-100 \%$ similar, whereas the Canadian isolate was only about $84-88 \%$ similar to the European group. The two strains may have diverged from each other in about 1900, at which time transfers of salmonids from North America to Europe (Rainbow trout) and from Europe to North America (sea run brown trout) were occurring. Both rainbow trout and
brown trout have been shown to be asymptomatic hosts of the virus. It is not known where the virus originated (Krøssoy et al. 2001).

Independent trials are underway in Canada to evaluate the efficacy of one of the ISA vaccines that is now widely used. The initial results have confirmed a significantly increased survival rate for fish that had been vaccinated.

### 2.4.2 Causes of fish farm escapes

The Working Group reviewed information on the reasons for the escape of farmed salmon from the British Columbia and East Coast North American salmon farming industries. This was considered relevant to the request to the Working Group (TOR 1.4) to assess possible reasons for the differences in the occurrence of escaped-farmed fish in fisheries and stocks in different areas. For these farmed salmon to find their way into fisheries or rivers, they first have to escape.

The British Columbia Fisheries Department has reviewed the causes of farm escape incidents that were reported to it by the salmon growers. Over the last five years the total number of reported escaped-farmed fish has stabilized at about $1 \%$ of the annual total salmon production ( $49,100 \mathrm{mt}$ in 1999 ; all species of salmon). On average, there were 5.2 reported escape incidents per year in 1996 - 2000 ( 26 events total reported over this time period). Escapes resulted from net failures ( $42 \%$ of the total; caused by predator attacks ( 6 of 11 net failures) and other factors), mechanical problems with cage systems or boats ( $4 \%$ of total), handling errors ( $39 \%$ ), and boat collisions with cages ( $15 \%$ ).

In the East Coast North American sea cage industry, the reporting of escapes has been imperfect, and the numbers of fish liberated is frequently uncertain. Six incidents could be documented between December 1999 and December 2000. The smallest escape of salmon was 3000 , and the largest $>100,000$. One event released 25,000 rainbow trout. Three of these six releases were storm related, one involved a boat collision, one was due to vandalism, and the cause for one is uncertain.

Escapes from individual fish farms in these two areas appear to predominantly result from inevitable human errors, and severe events like storms. While severe storms occur most frequently in autumn and winter, it will be difficult to predict when human error will occur. Consequently, the entry of farmed fish to the wild will retain a large degree of unpredictability. In addition, different salmon farming regions are characterized by different climates and operating conditions. Releases of farmed fish to the wild, and their occurrence in fisheries and rivers, will vary in both magnitude and frequency among these regions depending upon the severity of the conditions.

### 2.4.3 Differences in the occurrence of escaped farmed salmon in fisheries and stocks in different areas.

In 2000, about 627,000 tonnes of farmed salmon were produced in the Atlantic area, with Norway and Scotland accounting for the majority of production (see Section 2.2.1). In comparison, the total nominal landings of salmon in commercial fisheries in the north Atlantic in 2000 was about 2,800 tonnes. The catch included a relatively small proportion of salmon released as smolts for ranching, or for stock enhancement, and a proportion of escapees from fish farms. Salmon escape from fish farms at all life stages, they are caught in fisheries and enter freshwater to spawn (e.g. Hansen et al. 1987; Gausen \& Moen 1991; Webb \& Youngson, 1992; Youngson et al. 1997; Crozier 1998).

Farmed salmon are abundant in large numbers in Norwegian coastal commercial salmon fisheries. The proportion is lower in fjord and freshwater catches, but increases in spawning populations (Tables 2.4.3.1, 2.4.3.2). These differences have been suggested to be due to failure of the farmed salmon to home, and therefore the fish are not motivated to enter fjords and freshwater until later in the summer (Lund et al. 1991). Tagging experiments have shown that farmed salmon from Norway are caught in the Faroes fisheries (Hansen et al. 1987), and it has been shown that the incidence of escaped farmed salmon in this fishery can be high (Hansen et al. 1999). Estimates from the commercial fishery at West Greenland in 1991 and 1992 showed that the incidence of farmed fish was less than $1.5 \%$ (Hansen et al. 1997). Results from monitoring salmon fisheries and stocks in Scotland, Ireland and Northern Ireland have suggested a much lower proportion of farmed salmon (Webb \& Youngson 1992; Youngson et al. 1997; Crozier 1998; Tables 3.3.7.2-3.3.7.6) Fish farm escapees also occur in rivers in Canada and USA, particularly in areas with high density of farms. Estimates of the proportion of escaped farmed salmon in relation to nominal salmon catch in several countries are shown in Table 2.4.3.1 (ICES 2000/ACFM: 13). It should be noted however, that different methods used in the assessments as well as different geographical locations of the farms relative to salmon rivers could make it difficult to compare the figures between countries.

Analyses carried out in Norway have shown that the occurence of farmed salmon is highest in rivers close to areas with high density of fish farms. (Lund et al. 1994). In Ireland there have been 13 reported incidents between 1986 and 2000 involving 189,000 adults and 120,000 smolts escaping primarily from sites in the West, but also from sites in the North

West, South West and North. The relationship between the number of escapees in the declared catches and the reported salmon farm escapes is shown in Figure 2.4.3.1. The smolt escapes of any given year have been added to the following year to improve interpretation of the results on the assumption that the smolts could return in the following year. Although there are only 5 years data, the trend indicated would suggest that there is a relationship between the number of escapes and the number identified in the catch although these numbers are very low.

Wild salmon leave their home rivers as smolts in the spring and move quickly into oceanic areas (e.g. Holm et al. 1982). In the north east Atlantic areas results from smolt tagging experiments and post-smolt surveys have strongly indicated that ocean currents are the vectors that force the fish northwards (Jonsson et al. 1993; Shelton et al. 1997; Holst et al. 2000). Hatchery-reared salmon released as smolts in freshwater are thought to have a similar migratory pattern as wild salmon (Hansen et al. 1993). Hatchery smolts released on the coast tend to return to the same area from where they were released. (Carlin 1969; Sutterlin et al. 1979; Hansen et al. 1989), but apparently enter any river in that area to spawn.

Hansen \& Jonsson (1989; 1991) observed that when released tagged hatchery-reared salmon post-smolts kept in saltwater sequentially over one year, there was annual variation in both survival and homing precision, with poor survival of the groups released in late summer and autum, and poor homing precision of fish released in winter. Large salmon escaping early in the summer, a few months before spawning, tended to move northwards with the current, and when they were ready to spawn, they entered freshwater in that area. They did not appear to have a homing instinct (Hansen et al. 1987).

The returns from over 39,000 farmed fish experimentally tagged and released as smolts in Ireland 1984, 1985 and 1990 showed a very low rate of return in subsequent years. Compared to the return rates of between 3 and $13 \%$ on average for tagged smolts released for enhancement and for ranching, the rate of tag recovery was extremely low (less than $0.1 \%$ ) indicating that the survival rate of farmed fish which escape as smolts is very low.

An experiment carried out with large farmed salmon released from two farms on the Norwegian coast, Bersagel in south Norway and Meløy in mid Norway. The results were similar and supported the conclusions from studies cited above. Salmon that escape from fish farms in the autumn have lower survival rates than fish released in the winter/early spring (Table 2.4.3.2). The fish released from the two farms were recaptured in the sea, as well as in freshwater north of the site of release (Figure 2.4.3.2 and 2.4.3.3). Some of the fish released from the southern fish farm turned up in areas southeast of the site of release and entered freshwater in this area. Assuming that fish entering freshwater had made their final decision on where to spawn, it could be concluded that the farmed salmon used in the present experiment were not imprinted to any particular river or marine site, and could therefore be regarded as "homeless".

The distribution and direction of migration of the farmed salmon could be explained by transportation with currents (Figure. 2.4.3.4). If so, this may also explain why so few fish released in November and December were recovered. These fish would have been transported with the currents so far north that when they attained sexual maturity, they either were off route to detect freshwater, or they were simply lost in the cold Artic water. Fish that were released closer to maturity, might still have a higher probability to entering freshwater to spawn than groups released earlier, but the low recovery rates of these fish (less than $6 \%$ ) suggest that significant numbers of them were also lost.

Based on the current knowledge from the literature, the results from the tagging experiments, direction and speed of ocean currents, and from available information of the apparent low proportion of fish farm escapees in Ireland and Scotland relative to the production of farmed salmon, it is hypothetised that fish farm escapees from Faroes, Ireland and Scotland are transported with the currents, and fish that become sexually mature when they are relatively close to the coast enter Norwegian and Russian fisheries and salmon rivers. Under the same hypothesis some fish farm escapees from Ireland may enter fisheries and salmon rivers in N. Ireland and Scotland, some Irish and Scottish fish farm escapees may even turn up in Denmark and Sweden, and some Norwegian fish farm escapees may enter fisheries and rivers in Sweden, Denmark and Russia. It may be that a continous supply of fish farm escapees in the coastal current leads to a high proportion in Norwegian coastal salmon fisheries, although their survival are still low.

### 2.4.4 Causes of post-smolt mortality in the marine phase

## Possibility of by-catch of post-smolts in pelagic fisheries

Between 10 - 20 June 2000, special fishing experiments for post-smolts carried out in the Norwegian Sea yielded 268 post-smolts and 6 salmon in 14 tows during three consecutive days west and southwest of the Voeringplateau $\left(68^{\circ} 30^{\prime} \mathrm{N}\right.$ $-63^{\circ} \mathrm{N}$ and $1^{\circ} \mathrm{W}-5^{\circ} \mathrm{E}$ ) Table 3.9.1. Most of these fish were taken in three tows (170, 60, and 34 respectively, Figure 2.4.4.1). The CPUE at this particular cruise was 9 post-smolts per trawl hour, which is one of the highest recorded since 1990 (Table 3.9.1). Microtagged and Carlin-tagged fish occurred for the first time in the same hauls. In Norway no microtagging was carried out in 2000, indicating a south European origin of these fish, which supports the hypothesis
that south Norwegian fish and European fish are mixed on the feeding areas in the Norwegian Sea also at the post-smolt stage. These large catch numbers are of concern with respect to the potential impact of the mackerel fishery in the Norwegian Sea in June - August. There is overlap between the mackerel fishing areas and the anticipated northward migration routes for the post-smolts of south and central Europe and south- Norway (ICES 2000/ACFM13). The surface trawl method used by the Norwegian Research vessels resembles the commercial fishing method which also operates with a flotation on the trawl wings. However, the commercially used trawls are considerably deeper and longer, they are towed at higher speed, i.e. $\sim 5-6$ knots vs. 3-3.5 for the research ships, and the tows also last longer. The commercial trawlers thereby sweep much larger areas, and hence are likely to catch more post-smolts. So far it has not been possible to obtain detailed information on the methods used by the commercial ships, but the Norwegian Coastguards report a fleet of $25-30$ Russian and East European trawlers operating annually in the mackerel fishery in international area. Due to the assumed surface-near location of the post-smolts during migration and on their oceanic feeding grounds (Holm et al. 2000) the Working Group has previously recommended that ICES /NASCO should consider the advantages of commercial trawlers lowering the head ropes to a minimum of 5 m below the sea surface (ICES 1999/ACFM: 14) Furthermore the Working Group recommends that specific gear trials and extra observers are arranged for some periods during the pelagic fisheries in order to further investigate the possible impact on post-smolt survival.

## Salmon lice observations in selected Norwegian fjords and the Norwegian Sea

The status of salmon lice (Lepheoptheirus salmonis Krøyer) on seaward migrating post-smolts has been monitored by surface trawling in two southwest Norwegian fjords (Figure 2.4.4.2.) since 1998, i.e. since the live-fish sampling technique with the "Fish-lifter" (Holst and McDonald 2000) was introduced. In 2000, a special salmon survey was also carried out for the first time in four large north Norwegian fjords, and the Fish-lift technique was also used during a week in the Norwegian Sea (Table 3.9.1). This technique allows the majority of the fish to pass through the cod-end of the trawl with very little damage and loss of scales and external parasites (ICES 2000/ACFM:13).

The two southwest fjords were selected because they are different both hydrographically and in densities of fish farms, with the Nordfjord containing the largest number of net-pens. The northern fjords, again, represent areas with low (Altafjord) or no fish farms at all (Tana-/ Teno- and Neiden-/ Näätämöfjords) and large numbers of wild post-smolts (cf. Figure 2.4.4.3).

The northern post-smolt samples were infested with, on average, only 0.4 salmon lice per fish. The results of the analyses of salmon lice in post-smolt samples from the 1998-2000 captures in the Nordfjord and the Sognefjord are presented in Figure 2.4.4.3. The numbers have varied greatly between the years, especially in the Sognefjord, where the mean number of lice per fish has been over 30 the two last years. These particular outmigrating cohorts may therefore have been subjected to infestation rates surpassing even conservative estimates of lethal limits.

In the Nordfjord, which actually houses a high number of net-pens, the average number of lice per fish was relatively high in 1998-1999, while it was zero in the samples of 2000. This reflects a combination of an apparently recent entry of the post-smolts into the sea, and a thick layer of fresh water extending to the outlet of this fjord. Holst et al. (2001) report salinities of less than 10 ppm down to 4.5 m depth in the Nordfjord during the days the post-smolts were sampled. This may have protected the fish from infestation, thus underlining the possible importance for survival of the hydrography at the time of smolt passage through the fjords.

None of the samples analysed from the Norwegian Sea, carried more than 10 chalimus or older stages of lice per fish. This has been hypothesised to indicate that fish with high infestation rates either die, or lag behind the main cohorts of sea migrating post-smolts.

At present there is no data available to the Working Group that could enable correlation of the influence of the observed lice infestation rates on subsequent return rates of 1 SW or MSW salmon to the particular fjords. However, the high observed infestation rates are a matter of concern, which should be investigated in more detail.

### 2.4.5 Marine growth checks as evidence for sub-catchment population structuring.

The FRS, Freshwater Laboratory has routinely examined scale samples from adult Atlantic salmon (Salmo salar L.) returning to Scottish homewaters since the 1960s for the purposes of assessing age structure. As previously reported, scales from fish returning in 1997 showed a higher than previously recorded incidence of summer checks (ICES 1998/ACFM:15 and MacLean et al. 2000). This examination has been extended to incorporate salmon returning to Scottish home waters in 1998 and 1999.

The occurrence of a group of tightly-spaced circuli can be interpreted either as a winter annulus or as a summer growth check (Shearer, 1992). Recaptures of adult salmon previously tagged as emigrating North Esk smolts provide scale samples from fish whose sea age may also be derived from tagging records and thus where the presence of summer checks may be determined unambiguously. Between 1997 and 1999, 320 salmon which had been tagged as smolts on the North Esk were recaptured as returning adults in Scottish home water fisheries. Of these, 64 were identified as exhibiting growth checks on the marine zone of their scales. The sea-age of each fish was estimated both from scale samples and tagging records. In all cases, the ages derived from both methods corresponded and thus supported the interpretation of the growth checks as summer checks rather than as winter annulii.

The occurrence of summer checks on the scales of salmon returning to Scottish home waters in the years 1997-1999 was shown to be significantly greater than levels derived from the previous 35 -year period. There was no evidence that the incidence of checks varied between sexes. There was also no association between the presence of checks with either size at return or marine survival indices.

Where summer checks were identified on scales, the year when the check occurred was recorded, as was the relative position of the check within that year's marine growth zone. Examination of scale samples taken from the North Esk net \& coble fishery, which was the largest sample data set available, showed that checks were not distributed randomly over the marine zone. Three categories of summer check (1SW salmon returns and 2 SW returns with checks on the first or second summer at sea) were identified and their distribution of occurrence with respect to the growing season was analysed. The majority of checks tended to occur within a relatively narrow band within the third quarter of the marine zone. The proportion of salmon whose scales exhibited summer checks was highly variable both among years and sea age categories, but, in general, salmon showed a higher incidence of growth checks in their first year at sea than during their second year

The incidence of summer checks was also strongly related to the subsequent run-timing (the calendar month when fish returned to freshwater) of the adult fish. In particular, the incidence of summer checks in the first year of sea life was significantly associated with run-timing for each category of fish tested except 1SW salmon returning in 1999, when few checks were identified. Figure 2.4.5.1 shows the trends in incidence of summer checks with month of return to freshwater for the three categories of salmon. Seasonal patterns vary among groups but within each group, the pattern remained generally consistent between years.

The cause of the summer checks is unknown and the direct effects difficult to detect. Their relatively high incidence in recent years, however, may allow speculation on the mechanisms responsible for the observed patterns of association between groups of salmon in the ocean. While summer checks are present in scale samples taken from salmon returning to home waters throughout the sampling season the extent to which they occur varies both with adult run-time and sea age. These patterns of variation are relatively consistent among years suggesting that either different "run-timing groups" of salmon are differentially predisposed to the causal event or, salmon are not randomly mixed in the ocean and different groups follow, to some extent, different migration routes.

Radio tracking studies on a number of Scottish rivers show, within each sea age group, there is a relationship between the temporal pattern of return and the spatial distribution at spawning time (Anon, 1997, 1999; Laughton \& Smith, 1992; Smith et al., 1998, Walker \& Walker, 1991). Thus, the "run-timing groups" referred to above may be thought of as proxies for populations differentiated at a sub-catchment scale in freshwater. The pattern of association between individuals in the ocean as evidenced by the proportion whose scales show summer checks may thus reflect the subcatchment population structure found in rivers.

These observations may assist our understanding of recent trends in marine survival. Decreases in marine survival have been documented in the last decade throughout the north Atlantic at several monitored sites (ICES 2000/ACFM:13; Potter \& Crozier, 2000). Furthermore, differential rates of decline in different monthly components of the catch have also been documented (ICES 2000/ACFM:13). Early running spring salmon, in particular, appear to have declined most markedly (Youngson, 1995a) and the upper catchment populations associated with these runs of fish also mirror these declines (Youngson, 1995b; ICES 2000/ACFM:13). The structured variation in the incidence of summer checks between "run-timing groups" reported here provides an association through which differential trends in marine survival may occur as it demonstrates that coherent freshwater populations may encounter similar conditions in a patchy marine environment.

### 2.4.6 Estimates of $M$ at sea for Atlantic salmon

In the run-reconstruction models of the prefishery abundance (PFA) for the North American and Northeast Atlantic stock complexes, it is assumed that the natural mortality rate is $1 \%$ per month after the first year at sea. The assumed rate is from an analysis of weight and age data from the River Bush (U.K.) as developed by Doubleday et al. (1979)
(see below). This rate of natural mortality is used to calculate the number of fish immediately after the first winter, prior to the high seas fisheries, and between the high seas fisheries and returns to homewaters. If marine mortality rate is higher than previously assumed then its impact on assessments may be significant. In the time series of catches and returns used to estimate the PFA, there have been reductions in the level of sea fisheries such that presumably a smaller proportion of the estimated PFA consists of actual observed/harvested animals than was the case a decade ago (ICES CM 2000/ACFM:13). The concern is that the perception of reduced / declining abundance is in part an artifact of the model assumption about natural mortality during the second year at sea in terms of its assumed level and assumed constant rate over time. Two methods for estimating mortality at sea were reviewed, the inverse-weight method and the maturity schedule method.

## Method 1 - Inverse Weight Method

Ricker (1976) described a method for estimating the natural mortality rate based on the assumption that M decreases with increased size because marine natural mortality is assumed to be primarily the result of predation. Following on that approach, Doubleday et al. (1979) used the inverse weight hypothesis to estimate natural survival during the second year at sea based on catches, size-at-age, and return rates to the river and concluded that the natural mortality rate between Greenland and home waters (approx. 12 months) was between $3 \%$ and $12 \%$, i.e. about $1 \%$ per month. Lorenzen (1996) modelled the mortality of juvenile and adult fish as a power function of weight and using empirical observations of 113 species/stocks, derived parameter estimates for $M$ relative to weight. Based on these parameter values and using estimates of weight at age for River Bush salmon (tabled in Doubleday et al. 1979), the monthly mortality rate of Atlantic salmon in the second year of ocean life is about 3\% per month (Figure 2.4.6.1).

## Method 2 - Maturity Schedule Method

It is possible to estimate the sea survival rates of 1 SW and 2 SW salmon during the first and second years at sea by modeling the dynamics in the ocean using a simple life history model. Assuming that survival rates at age for males and females are similar, the model provides equations relating the survival rates and maturation profiles for 1SW and 2SW salmon.

$\mathrm{S}_{1} \mathrm{SW}_{\mathrm{M}},{\mathrm{S} 1 \mathrm{SW}_{\mathrm{F}}} \quad$ survival rates (relative to $\mathrm{N}_{0}$ ) of 1 SW salmon
$\mathrm{S}_{2} \mathrm{SW}_{\mathrm{M}}, \mathrm{S}_{2} \mathrm{SW}_{\mathrm{F}}$ survival rates (relative to $\mathrm{N}_{0}$ ) of 2 SW salmon

The four parameters to estimate and their constraints are:

$$
0=<S_{1}, S_{2}, \alpha_{M}, \alpha_{F}<=1
$$

The model was applied to data from three rivers:

1. Saint John River hatchery returns of age-1 smolts stocked at Mactaquac
2. LaHave River at Morgans Falls, wild smolts
3. Rivière de la Trinité (Québec) wild smolts

Sex ratios for the wild smolts were derived from sampling. Sex ratio for the age-1 smolts from Mactaquac were obtained from one year's sampling and assumed constant for the years analysed.

## Estimates of Marine Survival Rates

Survival rates during the first year at sea were low for the hatchery origin salmon of the Saint John River (range 1.4\% to $3.3 \%$ annual) but higher for wild smolts of de la Trinite River (range $1.5 \%$ to $8.5 \%$ ) (Figure 2.4.6.2). During the second year at sea, survival rates of the hatchery salmon ranged between $8 \%$ and $24 \%$ whereas the wild salmon survival rates in the second year at sea ranged between $17 \%$ and $79 \%$ (Figure 2.4.6.2). These survival rates are total survival rates after both natural and fishing mortality. Since 1992, most the sea fisheries have been closed or declining and the estimated survival rates can be considered equivalent to natural survival rates.

Survival rates during the first year have not responded to the closure of the fisheries in either of the stocks but survivals in the second year for de la Trinite salmon since the closure of the commercial fisheries are almost double the rates prior to the closure (Figure 2.4.6.2). Survival rates in both the first year and second year are better for wild smolts than hatchery smolts, and wild smolts from the northern stock (Trinite) are higher than those of the more southern LaHave River stock (Figure 2.4.6.3).

Mortality rates during the first and second years at sea are variable and since 1992 are high for both age groups. Based on the data from de la Trinite River, monthly Zs of between 0.02 and 0.15 have been estimated with the most frequently estimated value of 0.05 (Figure 2.4.6.4). Survival rates during the first year at sea have declined and in the 1990s remain as low or lower than those during the period of the 1980s when there were commercial fisheries. M therefore has increased over time.

The model results support the widely held view that the major source of mortality in the ocean occurs during the first year. They also provide evidence against the constant mortality rate assumptions used in the run-reconstruction model and for at least one wild stock of eastern Canada, monthly instantaneous mortality rates of $5 \%$ (ranging between $2 \%$ and $15 \%$ ) would seem more appropriate.

Although there appears to be increasing evidence of M being greater than $1 \%$ per month in the second year at sea and that M varies annually, the Working Group cautioned that only three rivers were evaluated and the data series on only one was longer than ten years. For this reason the revised values from preceding analysis have not been used in the forecast model for 2002. While an analysis of more rivers would be required to assess the among stock variability in the estimated survival rates and the representative level for the North American stock complex, it will be necessary to incorporate revised values for the forecast model in future as they become available. The Working Group recommended that further evaluation of the maturity schedule method be undertaken particularly as it relates to the sensitivities of the survival estimates to the sex ratio values of the smolts and the assumption of equal survival of male and female salmon.

### 2.4.7 Potential impact of climate change on juvenile salmon

Climate change has been identified as an important source of aquatic disturbance on a global scale and may alter species composition and dominance in aquatic ecosystems. Cold water ecosystems are particularly at risk and predictions from the climate change models for North America include:
from $1990-2100$, mean surface air temperature increases of $1.4-5.8^{\circ} \mathrm{C}$, with more rapid warming in the Northern regions of North America,
largest increases in air temperature in winter,
increased frequency and duration of summer hot spells (Hengeveld 1990),
increased water temperatures in the range of $2-5^{\circ} \mathrm{C}$ with maximum changes occurring in spring and fall,
advanced timing of snowmelt and spring runoff,
earlier start of a drier spring-summer season contributing to more extreme low $\mathbf{F}_{\text {low }}$ conditions (Manabe and Wetherald 1987).

Climate change has the potential to alter thermal regimes in aquatic environments, adversely affecting Atlantic salmon populations. Water temperature can affect survival, growth and behaviour of salmon in freshwater habitats. Juvenile Atlantic salmon begin feeding in the spring at water temperatures of $6-7^{\circ} \mathrm{C}$, and grow optimally at $16-19^{\circ} \mathrm{C}$. At water temperatures ranging from $22-24^{\circ} \mathrm{C}$, juvenile salmon have been observed to seek refuge from thermal stress. In some Atlantic salmon rivers in eastern Canada, juvenile salmon are already experiencing water temperatures approaching the upper lethal limit $\left(30^{\circ} \mathrm{C}\right)$.

The Working Group reviewed an analysis of the hydrological conditions and river temperatures in the Miramichi River over a 50 year time period and the associated variability in juvenile salmon size-at-age during 1971 to 1999.

Mean annual air temperature increased significantly, at a rate of $0.42^{*} \mathrm{C} /$ decade from $1970-1999$ with the warmest annual temperatures recorded in 1998 and 1999 (Figure 2.4.7.1). Mean air temperature in spring increased significantly, due to an increase of $0.58^{*} \mathrm{C} /$ decade in April ( $\mathrm{p}<0.011$ ). High temperatures were most frequently observed in 1999 and the frequency of high temperatures increased significantly in fall.

Mean summer water temperature was warmer in the Southwest Miramichi River than the Northwest Miramichi River, ranging from 12.3 to $15.3^{\circ} \mathrm{C}$. The warmest water temperatures in the 30 -year series were observed in 1999. Mean summer water temperature increased significantly from 1970 to 1999 , at a rate of $0.29^{\circ} \mathrm{C} /$ decade (p<0.037). The frequency of high water temperatures ranged from 22 days in 1986 to 114 days in 1999. The frequency of high water temperatures increased significantly during the parr growth season (early May to July 15), by approximately 4 days/decade ( $\mathrm{p}<0.041$ ).

The most significant change in discharge was observed in the timing of the spring snowmelt event which has shifted from April and May in the 1960s to March and April in the 1990s (Figure 2.4.7.2).

The range of mean annual fork length of Atlantic salmon fry was 4.0 to 5.4 cm (Figure 2.4.7.3). Mean annual fork length of $1+$ parr ranged from 7.6 to 9.1 cm , while size of 2+ parr ranged from 10.6 to 12.3 cm (Figure 2.4.7.4). Mean annual fork length of parr decreased significantly from 1970 to 1999 , at a rate of $0.18 \mathrm{~cm} /$ decade for $1+$ parr and 0.21 $\mathrm{cm} /$ decade for $2+\operatorname{parr}(\mathrm{p}<0.029)$.

In terms of the timing of the seasonal growth of parr, weight increased during spring and early summer, levelling off or decreasing slightly in late summer. Observed juvenile fish weight was consistently higher than that predicted from water temperature during May to August.

Annual and seasonal changes in meteorological and hydrological conditions were correlated with decreased fork length of juvenile Atlantic salmon. Fork length of parr was most strongly associated with maximum annual and spring air temperatures and mean spring water temperatures in the Southwest Miramichi River (Figure 2.4.7.4). Fork length of parr was also strongly associated with the frequency of high air temperatures during the parr growing season (May July 15) and the extreme high summer air temperatures.

Declines in fork length of juvenile Atlantic salmon parr ( $1+$ and $2+$ ) over the past 30 years suggest that conditions supporting growth have changed in the Northwest and Southwest Miramichi rivers. The functional model of growth relative to water temperature was a poor predictor of fish growth in the Miramichi River. The model tended to underestimate fish weight in spring and summer and overestimate in fall. For Atlantic salmon populations in the Miramichi, neither the theoretical functional model for maximum growth or growth potential index (as derived from the theoretical model) can be reliably used as predictors of fish growth in response to climate change.

Climate change is projected to have significant implications for aquatic ecosystems, altering thermal regimes and stream $\mathbf{F}_{\text {low }}$ conditions. The results of the analyses suggest that growth of juveniles in the Miramichi River are likely to be adversely affected by climate change, particularly during the spring months. Increases in air and water temperatures are expected to contribute to reduced size-at-age of juveniles with the potential effect of altering survival, age at smoltification, and ultimately sea survival

### 2.5 Compilation of Tag Releases and Finclip Data by ICES Member Countries in 2000

### 2.5.1 Compilation of tag releases and finclip data for 2000

Data on releases of tagged, fin-clipped, and marked salmon in 2000 were provided by the Working Group and are compiled as a separate report. A summary of Atlantic salmon marked in 2000 is given in Table 2.5.1. About 3.36 million salmon were marked in 2000, a decrease from the 4.43 million fish marked in 1999. The decrease was due largely to the reduced number of adipose fin clips. Primary marks are summarized in three classes: microtag (i.e., coded wire tag), external tag/mark, and adipose clips (without other external marks or fin clips. Secondary marks (primarily adipose clips on fish with coded wire tags) are also presented in the Annex. The adipose clip was the most used primary mark ( 2.35 million), with microtags ( 0.65 million) the next most used primary mark. Most marks were applied to hatchery-origin juveniles ( 3.30 million), while 44,115 wild juveniles and 16,150 adults were marked.



| 36e： | 94atal |  |  | Hexate |  |  | 7世木相 W | Hetart |  | $\begin{gathered} \text { Elum } \\ \text { cas } \end{gathered}$ |  |  |  |  |  |  | $\begin{gathered} =x_{1}^{2} \\ \frac{d y}{T} \end{gathered}$ |  <br>  <br> T |  |  |  |  |  | $\frac{\sqrt{4}}{\frac{4}{4}}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | $\frac{1 \times 2}{5}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 4 | 迷 | \％ |  |  | 3 |  | 4 | T | \％ | 0 | 7 | ＊ | 0 | 等 |  |  |  |  |  | ＊ | es | T |  |  |
| 1060： | ＂ | $\because$ | TEM | ＊ | $\cdots$ | $\cdots$ |  | ＊ | 160 |  | － | ＊ | W6 | ＊ | $\because$ | 100\％ | 16n | 33 | ＊ | 38 | 153 | 等落 | 48 | H4］ | 1 | 3\％ |
| 1084 | － | － | 13＊＊ | ＊ | $\cdots$ | ＊ | － | 教業 |  | － | － | w | － | － | 168\％ | w | 教 | 3 | 事業 | dt | 新复 | 3＊ | 事基 | 1 | ces |
| Wex | ＊ | ＊ | 3\％ | ＊ | ＊ | ＊ | $*$ | 迷 |  | ： | \％ | Haxay | ＊ | ＊ | 10x\％ | \％${ }^{\text {a }}$ | a | 44 | 迷 | 3is | 14x |  |  | 1 | 约：20 |
| 10．t | ＊ | ＊ | 12ex | ＊ | ＊ | ＊ | ＊ | 榣 |  | \％ | ＊ | Hex | ＊ | ＊ | 17040 | ，戓䢒 | 2＊ | 2 | 迷 | 3＊ | 䢒碞 | 輷㻃 |  | 1 | 敋延 |
| 100．0． | － | － | 粠 | － | － | ． | － | 政 |  | － | － | Wter | － | － | 教紼 | 淘 | 4 | \％ | 析 | 产为 | 12\％ | ＊＊ |  | 早 | 5400 |
| 130\％ | － | － | 3145 | － | － | － | － | 4x |  | － | － | 144 | － | － |  | ＊ | － | ， | 5x | W6． | 系的 | 血絽 | Inge | 1 |  |
| Onder | ＊ | $\cdots$ | \％ | ＊ | $\therefore$ | ＊ | ＊ | （1） | 寀 | ： | ＊ | 13＊ | ＊ | ＊ | 19\％10． | 5 | 4． | ＊ | 3x | ＊＊ | 103 ${ }^{\text {a }}$ | 緸爯 | H＊＊ | 4 | 荿数 |
| INers | \％ | $\pm$ | 3＊＊ | ＊ | ：＊ | $\pm$ | \％ | 1䋛 | $\underline{L}$ | $=$ | $=$ | Lucx | $=$ | ＊ | ${ }^{1}$ | 5 | cel | 3＊ | a | ＊ | 123 | 管道 | 2n？ | 1 | 鎄縉 |
| 150\％ | － | － | 321. | $\cdots$ | － | － | － | 185 | 1. | － | － | 1速发 | － | － | 146\％ | \％ | 䧸 | \％ | 趐 | 314 | 10\％ | 34\％ | ［474 | 1 |  |
|  | － | $\cdots$ | 3 | ． | $\cdots$ | ． | － | 184 | 考 | － | － | 1） | 枟竞 | ，${ }^{4}$ | 148\％ | （0） | ＋1． | 素 | \％${ }^{\text {\％}}$ | 48\％ | 3 ${ }^{3 / 8}$ |  | 珷等 | 1 | 紗冓 |
| \％ | 13，${ }^{\text {a }}$ | 3＊ |  | ＊ | － | s | $=$ | 363 | 14 | ＊ | ＊ | $1{ }^{1 / 4}$ | 娄娄 | 3＊ | 14＊ | 4， | 教 | 8 | 3 ${ }^{3}$ | 3＊ |  | \％ | 1989 | 1 | 察积 |
| 18\％ | 130．6 | 501 | 16＊＊ | ＊ | ＊ | ＊ | ＊ | 3＊＊ | \％ | \％ | ＊ | He＊＊ | 7 T |  | 12m | 4 | We | W | Hex | 3事 | 7教偖 | T） | Hatis | 1 | W9\％ |
| 160\％ | Hex | 縤 | 14＊＊ | ＊ | － | ，${ }^{\text {a }}$ | \％ |  | ＊ | 墄 | 13 Lk | W絲䋨 | 104\％ | \＄ | 1标虊 |  | 連 | 䓨 | 4， |  | twes |  | 基舜 | 1 |  |
| 10030 | 304 | 边 | 3䍃 | ＊ | ＊ | 年 | 業 | 130 | \％ | 緆 | Hext |  | 1） | 3640\％ | Hax | \％ | 4 | 等 | 4 |  | 12． | － | － $0_{0}$ | 管 | 變 |
| 19\％＊ | 14＊ | 30\％ | 家亚 | ． | ＊ | （0 | 等 | 龇娄 | 1 | 最 | 183＊ | 3138 | 1䍃 | －${ }^{\text {a }}$ | 1ex | W\％ | W | 3\％ |  |  |  | 3 \％ |  | 䋛 | 䓡細 |
| 13\％ | 137\％ | 51\％ |  | ＊ | ＊ | \％ | 3 | 1数 | 3 |  | 15x： | 3764 | 1mw | \％ | 1387 | 313 | \％ | 3 | $4{ }^{3}$ | 酸 | W\％ | 35． | 1＊ | 17 | 5xas |
| 1974 | 14 | ＊ | 8w | ＊ | － | \％ | 9 | 教 | 9 | 10\％ | 1＊＊＊＊ | 1 Wric | 1006 | （1） | 1＊＊＊ | ＊ | \％ | 30 | \％ | 14 | 32 | 4＊ | wnit | 路 |  |
| Heme | 薬尞 | will | \％ | ＊ | ＊ | 59 | 10 | 县䋛 | \％ | 称 | ＋1\％ | 1498 | 164\％ | 4，${ }^{\text {a }}$ | 14＊＊ | 为客 | 靽 | 18 | 建號 | ［妾 | ＊ | 数素 | \＃iva | ＋4 | W教 |
| 1946 | 1204 | 3x | 1考新 | ＊ | ＊ | 星 | 30 | 㹩参 | － | 14＊ | Heceis | 1200 | \％ | ane | 10tat | ＊ | 藓 | 10 | 刍数 | L繙 | 940 | 解番 | 1axay | ＊ |  |
| 1m\％ | 骨等 |  | 1487 | － | － | 3 | w | － | 6 | 13＊ |  | 102\％ | 1106 | mint | 1 16． | ＊＊＊ | ： | 13 | 政 | 9\％ | 3繻 | $4{ }^{4}$ | \％ |  | －6 |
| 109\％ | 1＊＊ | 等 | ＋6＊ | － | － | ，${ }^{1}$ | ＊ | 帾 | ＊ | 楼亚 | 34．4 | 雱 |  | ＋鞛 | 14＊＊ | ＊ | 4 | $1{ }^{\text {\％}}$ | 3w | 130 | 縤章 | 縤 | 1314 |  |  |
| 13＊ | W10．8 | （1） |  | ＊ | ＊ | 進 | W | 14＊ | W | 14． | 等缺 | \％ | 1趗 | 4n | 140， | ，${ }^{\text {ck }}$ | 2 | ＊ | 4＊ |  | － 4 | 變 | 1248 | $\cdots$ | 54 |
| 10\％ | 1838． | 316 | 13＊ | 4 | ＊ |  | 3 | 3 x | $1 \%$ | a | \％ | \％ | 瑗尞 | mam | 12＊ | sem | 峪 | \％ |  |  | 美䋛 | － |  | 6. | Ex\％ |
| 180＊ | 915 | 314．3 | 動＊ | ＊ | \％ | 动等 | 36000 | 迹 | 景缶 | 新路 | 1\％＊ | neke | ＊27 | ＊＊ | 14＊\％ | \％ | 4． | 緆 | 城复 | 迷数 | －${ }^{2}$ | ＊${ }^{3}$ |  | 4 | Y蒳 |
| 380 | \％ | ＊ | 㙖書 | \％ | 0 | － | ＊ | 36 | \＄＊ | amb |  | \％ | \％ | \％ | 160 | 如家 | 堂 | ＋ | 34 | 絽 | ＊ | 等慗 | 364 | 書 |  |
| 108 |  | 5 | Hes | 38 | 4 | 4 | 这 | 3630 | 菜 | amo | 14080 | 1，wh | 382 | （2） | 184． | －${ }^{5}$ | 雨 | 4 | $3{ }^{4}$ | \％ | 3新 | 130 | 整尞 | 垩 | －6．s． |
| T\％ | 789 | 7 | 13＊＊＊ | 24 | $1{ }^{1}$ | \％ | \％ | 3 | 5 | 14 | 3 ${ }^{\text {che }}$ | 170 | 1604 | 易数 | 13＊＊ | new | 2 | ＊＊ | m | low | 3＊ | 32\％ | Ex． | $1{ }^{1 / 4}$ | TMex |
| 1685 | 35 | 新逪 | 1\％＊＊ | 3 | ＋ | 4 | 数 | 1310 | 唓 | 糕 | 114 |  | 景 | － | 13060 |  | 14． | a | \％ | ＊ | 54． | 建 | 䇣等 | 13， | 納䋛 |
| 10080 | 家 |  | 410 | 新 | $\rho$ | 4 | \％ |  | 皿尞 | 14． | H |  | 管等 | 4 | 10才6 |  | \％ | － | 楼 | 14 |  | 3緒 | 等 | 崖类 |  |
| 10＊ | W | 緒 | 115\％ | 䵭 | 䓣 | 䍃 | 4 | lan | 4 | 樓 | 3\％ | May | － | （鳞 |  | 5 | \％ | \％ | － | 動嘘 | －${ }^{\text {a }}$ | 4 | 䜌 | 4 | \％ |
| 10\％ | 朝 | （ex | ＊11 | 4 | － | （\％） | s＊ | 蝛 | 运 | － | － | ＊ |  | 綅 | \％＊ | ＊ | \％ | 3 | ＊ | 50 | ＊2 |  | 粦 | 者迷 | ＋60 |
| 198． | 30 | Sil | 11 | W | 1 | ＊ | $3{ }^{3}$ | 46 | ＊ | － | － |  | 38 | 44＊ | 影絡 | 速教 | ＊＊ | 绪 | \％${ }^{\text {a }}$ | 縉 |  |  | －${ }^{\text {\％}}$ | 奚 | \％ |
| 或楽缶 | W | 10\％ | 53x | 4 | T | 管 | w | 㸻数 | ＊ | ＊ | ＝ |  | \％ | 3xal | 3 | 沓 | \％ | ＊ | 12 ${ }^{2}$ | W | 3 38． | 3＊ |  | 4 | 楽䋨 |
| 180 | 214 | 1200 | 37 | 5 | 17 | T | \％ | 100 | \％ | ＊ | n | 3 | 61. | 3 Lz | 32 | 38 | \％ | \％ | 䋹 | － | 3部 | za |  | 36 | 34x |
| $1{ }^{\text {cken }}$ | 314 | 140\％ | 器数 | 楼 | 1. | ＊ | 䗇 | 3，${ }_{\text {a }}$ | we | － | － | 絞䢒 | 䜌 | 4tis | － | 綡 | W | 4 | 隹事 | ＊ | \％${ }^{\text {a }}$ | 2t | －${ }^{\text {W }}$ | － | 絲䜌 |
| 14048 | 1430 | 14＊ | 35 | 等 | 1 | 4 | \％ | ，${ }^{\text {cke }}$ | 䊾䍃 | － | \％ | 30\％ | 300 | $3 \times$ | 歯数 | 䌯 | ＊ | 3 | －${ }^{4}$ | 縣 | ask | \％${ }^{\text {\％}}$ | 號 | － | 3 SW |
| 1456 |  | 13＊ | 这串 | － | 31. | 擞事 |  |  | 等运 | ： | $\because$ | $6_{6} 8$ | 3 ${ }^{3}$ | 20x | 需 | 3震 | \％ | 3 |  | 39 | 嘘 | 䜌 | x ${ }^{2}$ | \％ | 3，${ }^{\text {a }}$ |
| 189\％ | 120 | 10\％ | 230 | 2 | 䌊 | 4 | 者 | 10\％ | ＊${ }^{*}$ | ＂ | － | \％ | 遙 | zat | 53\％ | 314 | 3 | 閣 | $1{ }^{4}$ | 坥 | 锁 | 洓要事 | 橖 | ＊ | 2mat |
| 124 | \％ | 4 | $1{ }^{1}$ | $\pm$ | ＋ | － | 4 | 24 | ， | － | － | cos | 織 | － 4 | 教都 | 13 | 4 | $1{ }^{4}$ | 14 | 3 | \％ |  | 縉 | － | 4 4 |
| 18\％ | \％ | \％ | 1管 | 3 | ＊ | a | 者 | 30 | ＊ | － | － | 等奚 | － | 313等 | 期这 | \％ | 6 | \％ | 鯺 | 號 | ＋4 | 䜌 | W | － |  |
| 3xa | ca | 8 | 13 | 3 | 35 | 93 | 17 | 3＊ | $\pm$ | ＝ | $=$ | 6s | 63 | 301 | 11 ＊ | 120 | \％ min $^{\text {a }}$ | 3 | 21\％ | \％ | ［18） | 0 | － | $=$ | zue |
| max |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 10\％ | 12 | 10 | 处粦 | 3 | 3 | ＊ | w | 14＊ | 123 | － | － | \％ 6 |  | 喊迷 |  | 枟县 | \％ | 建 | 12 | \％ |  | 20 | \％䜌 | $=$ |  |
| 3n＊006 | 景前 | 1＊ | 发䒨 | a | 0 | 54 | 13 | 4＊＊ | ＊＊ | － | \％ | 迷妾 | 盛全 | 30 |  |  | \％ | 34 | 立管 | 8 | ，縲 | 3 | 4 | 4 |  |










Table 2.1.1.3 Reported catch of SALMON in numbers and weight in tonnes (round fresh weight). Catches reported for 2000 may be provisional. Methods used for estimating age composition given in footnotes.


Table 2.1.1.3 continued


Table 2.1.1.3 continued


Table 2.1.1.3 continued


Table 2.1.1.3 continued

| Country | Year | 1SW |  | 2SW |  | 3SW |  |  | 4SW |  |  | 5SW |  |  | MSW ${ }^{1}$ |  | PS |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | No. | Wt | No. | Wt | No. |  | Wt | No. |  | Wt | No. |  | Wt | No. | Wt | No. | Wt | No. | Wt |
| $\begin{aligned} & \hline \mathrm{UK} \\ & \text { (Scotland) } \end{aligned}$ | 1982 | 208,061 | 416 | - |  |  | - |  |  |  |  |  |  |  | 128,242 | 596 |  |  | 336,3032 | 1,092 |
|  | 1983 | 209,617 | 549 | - |  |  | - |  |  |  |  |  | - |  | 145,961 | 672 | - |  | 320,578 | 1,221 |
|  | 1984 | 213,079 | 509 | - |  |  | - |  |  |  |  |  | - |  | 107,213 | 504 | - |  | 230,292 | 1,013 |
|  | 1985 | 158,012 | 399 | - |  |  | - |  |  |  |  |  | - |  | 114,648 | 514 | - |  | 272,660 | 913 |
|  | 1986 | 202,861 | 526 | - |  |  | - |  |  |  |  |  | - |  | 148,398 | 745 | - |  | 351,259 | 1,271 |
|  | 1987 | 164,785 | 419 | - |  |  | - |  |  |  |  |  | - |  | 103,994 | 503 | - |  | 268,779 | 922 |
|  | 1988 | 149,098 | 381 | - |  |  | - |  |  |  |  |  | - |  | 112,162 | 501 | - |  | 261,260 | 882 |
|  | 1989 | 174,941 | 431 | - |  |  | - |  |  |  |  |  | - |  | 103,886 | 464 | - |  | 278,827 | 895 |
|  | 1990 | 81,094 | 201 | - |  |  | - |  |  |  |  |  | - |  | 87,924 | 423 | - |  | 169,018 | 624 |
|  | 1991 | 73,608 | 177 | - |  |  | - |  |  |  |  |  | - |  | 65,193 | 285 | - |  | 138,801 | 462 |
|  | 1992 | 101,676 | 238 | - |  |  | - |  |  |  |  |  | - |  | 82,841 | 361 | - |  | 184,517 | 600 |
|  | 1993 | 94,517 | 227 | - |  |  | - |  |  |  |  |  | - |  | 71,726 | 320 | - |  | 166,243 | 547 |
|  | 1994 | 99,459 | 248 | - |  |  | - |  |  |  |  |  | - |  | 85,404 | 400 | - |  | 184,863 | 649 |
|  | 1995 | 89,921 | 224 | - |  |  | - |  |  |  |  |  | - |  | 78,452 | 364 | - |  | 168,373 | 588 |
|  | 1996 | 66,413 | 160 | - |  |  | - |  |  |  |  |  | - |  | 57,920 | 267 | - |  | 124,333 | 427 |
|  | 1997 | 46,872 | 114 | - |  |  | - |  |  |  |  |  | - |  | 40,427 | 182 | - |  | 87,299 | 296 |
|  | 1998 | 53,447 | 121 | - |  |  | - |  |  |  |  |  | - |  | 39,248 | 162 | - |  | 92,695 | 283 |
|  | 1999 | 25,183 | 57 | - |  |  | - |  |  |  |  |  | - |  | 30,651 | 142 | - |  | 55,834 | 199 |
|  | 2000 | 26,896 | 66 |  |  |  | - |  |  |  |  |  |  |  | 29,774 | 126 | - |  | 56,671 | 192 |
| USA | 1982 | 33 |  | 1,206 |  |  | 5 |  |  |  |  |  | - |  |  | - | 21 |  | 1,265 | 6.4 |
|  | 1983 | 26 | - | 314 |  |  | 2 |  |  | - |  |  | - |  | - | - | 6 |  | 348 | 1.3 |
|  | 1984 | 50 | - | 545 |  |  | 2 |  |  | - |  |  | - |  | - | - | 12 |  | 609 | 2.2 |
|  | 1985 | 23 | - | 528 |  |  | 2 |  |  | - |  |  | - |  | - | - | 13 |  | 557 | 2.1 |
|  | 1986 | 76 | - | 482 |  |  | 2 |  |  | - |  |  | - |  | - | - | 3 |  | 541 | 1.9 |
|  | 1987 | 33 | - | 229 |  |  | 10 |  |  | - |  |  | - |  | - | - | 10 |  | 282 | 1.2 |
|  | 1988 | 49 | - | 203 |  |  | 3 |  |  | - |  |  | - |  | - | - | 4 |  | 259 | 0.9 |
|  | 1989 | 157 | 0.3 | 325 |  |  | 2 |  |  | - |  |  | - |  | - | - | 3 |  | 487 | 1.7 |
|  | 1990 | 52 | 0.1 | 562 |  |  | 12 |  |  | - |  |  | - |  | - | - | 16 |  | 642 | 2.4 |
|  | 1991 | 48 | 0.1 | 185 |  |  | 1 |  |  | - |  |  | - |  | - | - | 4 |  | 238 | 0.8 |
|  | 1992 | 54 | 0.1 | 138 |  |  | 1 |  |  | - |  |  | - |  | - | - | - |  | 193 | 0.7 |
|  | 1993 | 17 |  | 133 |  |  | - |  |  | - |  |  | - |  | - | - | 2 |  | 152 | 0.6 |
|  | 1994 | 12 | - | 0 |  |  | - |  |  | - |  |  | - |  | - | - | - |  | 12 | 0 |
|  | 1995 | - | - | - |  |  | - |  |  | - |  |  | - |  | - | - | - |  | 0 | 0 |
|  | 1996 | - | - | - |  |  | - |  |  | - |  |  | - |  | - | - | - |  | 0 | 0 |
|  | 1997 | - | - | - |  |  | - |  |  | - |  |  | - |  | - | - | - |  | 0 | 0 |
|  | 1998 | - | - | - |  |  | - |  |  | - |  |  | - |  | - | - | - |  | 0 | 0 |
|  | 1999 | - | - | - |  |  | - |  |  | - |  |  | - |  | - | - | - |  | 0 | 0 |
|  | 2000 |  |  | - |  |  | - |  |  | - |  |  | - |  |  |  | - |  | 0 | 0 |

Table 2.1.1.3 continued


MSW includes all sea ages $>1$, when this cannot be broken down.
Different methods are used to separate 1SW and MSW salmon in different countries:

- Scale reading: Faroe Islands, Finland (1996 onwards), France, Russia, UK (England and Wales), USA and West Greenland.
- Size (split weight/length): Canada ( 2.7 kg for nets; 63 cm for rods), Finland up until 1995 ( 3 kg ), Iceland (various splits used at different times and places),
- Norway ( 3 kg ), UK (Scotland) ( 3 kg in some places and 3.7 kg in others). All countries except Scotland report no problems with using weight to catergorise - catches into sea age classes.

In Scotland, misclassification may be very high in some years.
In Norway, catches shown as 3SW refer to salmon of 3SW or greater.
2. Data for 1993-98 altered from previous reports to take account of catch \& release

Table 2.1.1.4 The weight (tonnes round fresh weight) and proportion (\%) of the nominal catch by country taken in coastal, estuarine and riverine fisheries.

| Country | Year | Catch |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Coast |  | Estuary |  | River |  | Total <br> Weight |
|  |  | Weight | \% | Weight | \% | Weight | \% |  |
| Canada | 1999 | 7 | 5 | 38 | 25 | 105 | 70 | 150 |
|  | 2000 | 11 | 7 | 22 | 15 | 117 | 78 | 150 |
| Finland | 1995 | 0 | 0 | 0 | 0 | 48 | 100 | 48 |
|  | 1996 | 0 | 0 | 0 | 0 | 44 | 100 | 44 |
|  | 1997 | 0 | 0 | 0 | 0 | 45 | 100 | 45 |
|  | 1998 | 0 | 0 | 0 | 0 | 48 | 100 | 48 |
|  | 1999 | 0 | 0 | 0 | 0 | 63 | 100 | 63 |
|  | 2000 | 0 | 0 | 0 | 0 | 95 | 100 | 95 |
| $\text { France }{ }^{1}$ | 1995 | - | - | 2 | 20 | 8 | 80 | 10 |
|  | 1996 | - | - | 4 | 31 | 9 | 69 | 13 |
|  | 1997 | - | - | 3 | 38 | 5 | 63 | 8 |
|  | 1998 | 1 | 13 | 2 | 25 | 5 | 63 | 8 |
|  | 1999 | 0 | 0 | 4 | 35 | 7 | 65 | 11 |
|  | 2000 | 0 | 4 | 4 | 35 | 7 | 61 | 11 |
| Iceland | 1995 | 20 | 13 | 0 | 0 | 130 | 87 | 150 |
|  | 1996 | 11 | 9 | 0 | 0 | 111 | 91 | 122 |
|  | 1997 | 0 | 0 | 0 | 0 | 106 | 100 | 106 |
|  | 1998 | 0 | 0 | 0 | 0 | 130 | 100 | 130 |
|  | 1999 | 0 | 0 | 0 | 0 | 119 | 100 | 119 |
|  | 2000 | 0 | 0 | 0 | 0 | 82 | 100 | 82 |
| Ireland | 1995 | 566 | 72 | 140 | 18 | 84 | 11 | 790 |
|  | 1996 | 440 | 64 | 134 | 20 | 110 | 16 | 684 |
|  | 1997 | 380 | 67 | 100 | 18 | 91 | 16 | 571 |
|  | 1998 | 433 | 69 | 92 | 15 | 99 | 16 | 624 |
|  | 1999 | 335 | 65 | 83 | 16 | 97 | 19 | 515 |
|  | 2000 | 440 | 71 | 79 | 13 | 102 | 16 | 621 |
| Noway | 1995 | 515 | 61 | 0 | 0 | 325 | 39 | 840 |
|  | 1996 | 520 | 66 | 0 | 0 | 267 | 34 | 787 |
|  | 1997 | 394 | 63 | 0 | 0 | 235 | 37 | 629 |
|  | 1998 | 410 | 55 | 0 | 0 | 331 | 45 | 741 |
|  | 1999 | 483 | 60 | 0 | 0 | 327 | 40 | 810 |
|  | 2000 | 619 | 53 | 0 | 0 | 557 | 47 | 1176 |
| Russia | 1995 | 43 | 33 | 9 | 7 | 77 | 60 | 128 |
|  | 1996 | 64 | 49 | 21 | 16 | 46 | 35 | 131 |
|  | 1997 | 63 | 57 | 17 | 15 | 32 | 28 | 111 |
|  | 1998 | 55 | 42 | 2 | 2 | 74 | 56 | 131 |
|  | 1999 | 48 | 47 | 2 | 2 | 52 | 51 | 102 |
|  | 2000 | 64 | 52 | 15 | 12 | 45 | 36 | 124 |


| Country | Year | Catch |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Coast |  | Estuary |  | River |  | Total Weight |
|  |  | Weight | \% | Weight | $\%$ | Weight | \% |  |
| Spain | 1995 | 0 | 0 | 0 | 0 | 9 | 100 | 9 |
|  | 1996 | 0 | 0 | 0 | 0 | 7 | 100 | 7 |
|  | 1997 | 0 | 0 | 0 | 0 | 4 | 100 | 4 |
|  | 1998 | 0 | 0 | 0 | 0 | 4 | 100 | 4 |
|  | 1999 | 0 | 0 | 0 | 0 | 6 | 100 | 6 |
|  | 2000 | n/a | - | nia | - | nua | - | nua |
| Sweden | 1995 | 24 | 65 | 0 | 0 | 13 | 35 | 37 |
|  | 1996 | 19 | 58 | 0 | 0 | 14 | 42 | 33 |
|  | 1997 | 10 | 56 | 0 | 0 | 8 | 44 | 18 |
|  | 1998 | 5 | 33 | 0 | 0 | 10 | 67 | 15 |
|  | 1999 | 5 | 31 | 0 | 0 | 11 | 69 | 16 |
|  | 2000 | 10 | 30 | 0 | 0 | 23 | 70 | 33 |
| UK | 1995 | 200 | 68 | 45 | 15 | 49 | 17 | 294 |
| England \& | 1996 | 83 | 45 | 42 | 23 | 58 | 32 | 183 |
| Wales | 1997 | 81 | 57 | 27 | 19 | 35 | 24 | 143 |
|  | 1998 | 65 | 53 | 19 | 16 | 38 | 31 | 122 |
|  | 1999 | 101 | 67 | 23 | 15 | 26 | 17 | 150 |
|  | 2000 | 152 | 71 | 25 | 12 | 36 | 17 | 213 |
| UK (N. Ireland) ${ }^{2}$ | 1999 | 44 | 83 | 9 | 17 | 0 | 0 | 53 |
|  | 2000 | 63 | 82 | 14 | 18 | 0 | 0 | 77 |
| UK | 1995 | 201 | 34 | 105 | 18 | 282 | 48 | 588 |
| Scotland | 1996 | 129 | 30 | 80 | 19 | 218 | 51 | 427 |
|  | 1997 | 79 | 27 | 33 | 11 | 184 | 62 | 296 |
|  | 1998 | 60 | 21 | 28 | 10 | 195 | 69 | 283 |
|  | 1999 | 35 | 18 | 23 | 12 | 141 | 71 | 199 |
|  | 2000 | 30 | 16 | 24 | 12 | 139 | 72 | 193 |
| Totals |  |  |  |  |  |  |  |  |
| North East Atlantic ${ }^{3}$ | 2000 | 1386 | 53 | 161 | 6 | 1086 | 41 | 2633 |
| North America ${ }^{4}$ | 2000 | 13 | 9 | 22 | 14 | 117 | 77 | 152 |

'An illegal net fishery operated from 1995 to 1998 , catch unknowm in the first 3 years but
thought to be increasing. Fishery ceased in 1999
${ }^{2}$ no nominal catch data is collected for river fisheries in UK (NI)
${ }^{3}$ data not available from Dermark \& Spain
${ }^{4}$ inchudes Canada \& St Pierre et Miquelon

Table 2.1.2.1 Numbers of fish caught and released in rod fisheries along with the $\%$ of the total rod catch (released + retained) for countries in the North Atlantic where records are available, 1991-2000.

| Year | Canada ${ }^{1}$ |  | Iceland |  | Russia |  | UK(E\&W) |  | UK(Scot) |  | USA |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total | $\begin{gathered} \% \text { of total } \\ \text { rod } \\ \text { catch } \end{gathered}$ | Total | $\begin{gathered} \% \text { of total } \\ \text { rod } \\ \text { catch } \end{gathered}$ | Total | \% of total <br> rod catch | Total | $\begin{aligned} & \text { \% of total } \\ & \text { rod } \\ & \text { catch } \end{aligned}$ | Total | $\begin{gathered} \% \text { of total } \\ \text { rod } \\ \text { catch } \end{gathered}$ | Total | $\begin{gathered} \% \text { of total } \\ \text { rod } \\ \text { catch } \end{gathered}$ |
| 1991 |  |  |  |  | 3,211 | 51 |  |  |  |  | 239 | 50 |
| 1992 | 46,450 | 34 |  |  | 10,120 | 73 |  |  |  |  | 407 | 67 |
| 1993 | 53,849 | 41 |  |  | 11,246 | 82 | 1,448 | 10 |  |  | 507 | 77 |
| 1994 | 45,804 | 39 |  |  | 12,056 | 83 | 3,227 | 13 | 6,595 | 8 | 249 | 95 |
| 1995 | 31,211 | 36 |  |  | 11,904 | 84 | 3,189 | 20 | 12,133 | 14 | 370 | 100 |
| 1996 | 36,934 | 33 | 669 | 2 | 10,745 | 73 | 3,428 | 20 | 10,409 | 15 | 542 | 100 |
| 1997 | 48,387 | 49 | 1,558 | 5 | 14,823 | 87 | 3,132 | 24 | 10,906 | 18 | 333 | 100 |
| 1998 | 56,860 | 52 | 2,826 | 7 | 12,776 | 81 | 5,365 | 31 | 13,455 | 18 | 273 | 100 |
| 1999 | 49,268 | 50 | 3,051 | 10 | 11,450 | 77 | 5,447 | 44 | 14,839 | 28 | 211 | 100 |
| $2000^{2}$ | 49,737 | 53 | 2,691 | 10 | 12,914 | 74 | 7,355 | 42 | 19,991 | 34 | - | - |

1. Figures for 1992 to 1996 are minimal estimates as not all areas have
reported catch and release.
2. Figures for 2000 are provisional.

Table 2.1.3.1 Estimates of unreported catches by various methods in tonnes within

|  | national EEZs in the North-East Atlantic, North American |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | and West Greenland Commissions of NASCO, 1986-2000. |  |  |  |
| Year | North-East | North-American | West | Total |
|  | Atlantic |  | Greenland |  |
| 1987 | 2,554 | 234 | - | 2,788 |
| 1988 | 3,087 | 161 | - | 3,248 |
| 1989 | 2,103 | 174 | - | 2,277 |
| 1990 | 1,779 | 111 | - | 1,890 |
| 1991 | 1,555 | 127 | - | 1,682 |
| 1992 | 1,825 | 137 | - | 1,962 |
| 1993 | 1,471 | 161 | $<12$ | 1,644 |
| 1994 | 1,157 | 107 | $<12$ | 1,276 |
| 1995 | 942 | 98 | 20 | 1,060 |
| 1996 | 947 | 156 | 20 | 1,123 |
| 1997 | 732 | 90 | 5 | 827 |
| 1998 | 1,108 | 91 | 11 | 1,210 |
| 1999 | 887 | 133 | 12,5 | 1,032 |
| 2000 | 1,135 | 124 | 10 | 1,269 |
| Mean |  |  |  |  |
| 1995-1999 | 923 | 114 | $<14$ | 1050 |

Table 2.1.3.2 Estimates of unreported catches by various methods in tonnes by country within national EEZs in the North-East Atlantic, North America and West Greenland Commissions of NASCO, 2000, (NA = not available).

| 2000 Commission Area | Country | Unreported Catch t | Unreported as \% of Total <br> North Atlantic Catch <br> (Unreported + Reported) | Unreported as \% of Total <br> National Catch <br> (Unreported + Reported) |
| :---: | :---: | :---: | :---: | :---: |
| NEAC | Faroes | $<1$ | - | - |
| NEAC | Finland | 25 | 0.6 | 21 |
| NEAC | Iceland | 2 | 0.0 | 2 |
| NEAC | Ireland | 132 | 3.2 | 18 |
| NEAC | Norway | 633 | 15.5 | 35 |
| NEAC | Russia | 250 | 6.1 | 67 |
| NEAC | Sweden | 4 | 0.1 | 11 |
| NEAC | UK (E \& W) | 38 | 0.9 | 15 |
| NEAC | UK (N.Ireland) | 8 | 0.2 | 9 |
| NEAC | UK (Scotland) | 44 | 1.1 | 19 |
| NAC | Canada | 124 | 3.0 | 45 |
| NAC | USA | 0 | 0.0 | 0 |
| WGC | West Greenland | 10 | 0.2 | 32 |
|  | Total Unreported Catch | 1269 | 31.1 |  |
|  | Total Reported Catch of North Atlantic salmon | 2814 |  |  |

Table 2.2.1.1 Production of farmed salmon in the North Atlantic area and in areas other than the North Atlantic (in tonnes round fresh weight), 1980-2000.


Table 2.2.2.1 Production of ranched salmon in the North Atlantic (tonnes round fresh weight) as harvested at ranching facilities, 1980-2000.

| Year | Iceland <br> commercial <br> ranching | Ireland ${ }^{1}$ | UK(N.Ireland) <br> River <br> Bush ${ }^{1}$ | Norway <br> various <br> facilities ${ }^{1}$ | Total <br> production |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 8.0 |  |  | 0.0 | 8.0 |
| 1981 | 16.0 |  |  | 0.0 | 16.0 |
| 1982 | 17.0 |  |  |  | 17.0 |
| 1983 | 32.0 |  |  |  | 32.0 |
| 1984 | 20.0 |  |  |  | 20.0 |
| 1985 | 55.0 | 17.5 | 17.0 |  | 89.5 |
| 1986 | 59.0 | 22.9 | 22.0 |  | 103.9 |
| 1987 | 40.0 | 6.4 | 7.0 |  | 53.4 |
| 1988 | 180.0 | 11.5 | 12.0 | 4.0 | 207.5 |
| 1989 | 136.0 | 16.3 | 17.0 | 3.0 | 172.3 |
| 1990 | 280.0 | 5.7 | 5.0 | 6.0 | 296.7 |
| 1991 | 345.0 | 3.6 | 4.0 | 5.0 | 357.6 |
| 1992 | 460.0 | 9.4 | 11.0 | 10.0 | 490.4 |
| 1993 | 496.0 | 9.7 | 8.0 | 11.0 | 524.7 |
| 1994 | 308.0 | 15.2 | 0.4 | 9.5 | 333.1 |
| 1995 | 298.0 | 16.8 | 1.2 | 2.0 | 318.0 |
| 1996 | 239.0 | 18.5 | 3.0 | 8.0 | 268.5 |
| 1997 | 50.0 | 4.1 | 2.8 | 2.0 | 58.9 |
| 1998 | 34.0 | 9.6 | 1.0 | 1.0 | 45.6 |
| 1999 | 26.0 | 4.3 | 1.4 | 1.0 | 32.7 |
| 2000 | 2.0 | 4.6 | 3.5 | 1.0 | 11.1 |
| Mean |  |  |  |  |  |
| $1995-99$ | 129 | 11 | 2 | 3 | 145 |

${ }^{1}$ Total yield in homewater fisheries and rivers.

Table 2.4.3.1 Proportion of escaped farmed salmon in relation to nominal salmon catch (ICES 2000).

| Year | Norway | Faroes | Ireland | N. Ireland | Scotland |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1989 | 22 |  |  |  |  |
| 1990 | 23 |  |  |  |  |
| 1991 | 22 | 42 | 0.4 | 1.8 | 3.0 |
| 1992 | 23 | 34 | 0.6 | 1.2 | 5.2 |
| 1993 | 23 | 27 | 0.4 | 0.2 | 5.7 |
| 1994 | 21 | 17 | 0.3 | 0.5 | 0.8 |
| 1995 | 22 | 20 | 0.1 | 1.8 | 0.3 |
| 1996 | 28 | 20 | 0.2 |  | 0.2 |
| 1997 | 31 |  | 0.2 | 0.1 | 0.3 |
| 1998 | 28 |  | 0.3 | 0.0 | 0.4 |
| 1999 | 24 |  | 0.4 | 1.3 | 0.5 |

Table 2.4.3.2. Number of recoveries and recapture rates of farmed salmon released from two fish farms in Norway.

| Meløy |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Date of release | No in marine <br> fisheries | No in <br> freshwater | Total Number | Recapture <br> Rate (\%) |
| 03.11 .93 | 1 | 0 | 1 | 0.2 |
| 16.12 .93 | 4 | 0 | 4 | 0.8 |
| 02.02 .94 | 8 | 1 | 9 | 1.9 |
| 23.03 .94 | 21 | 6 | 27 | 5.5 |
| Bersagel |  |  |  |  |
| 12.11 .93 | 1 | 0 | 1 | 0.2 |
| 17.12 .93 | 2 | 5 | 7 | 1.4 |
| 18.02 .94 | 5 | 1 | 6 | 1.3 |
| 24.03 .94 | 12 | 6 | 18 | 3.8 |
| 25.04 .94 | 17 | 5 | 22 | 4.5 |

Table 2.5.1 Summary of Atlantic salmon tagged and marked in 2000. 'Hatchery' anc 'Wild' refer to smolts or parr; 'Adult' refers to wild and hatchery fish. Data from Belgium, France, and Spain were not available. Fish were not tagged in Finland.

| Country | Origin | Primary Tag or Mark |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Microtag | External mark | Adipose clip |  |
| Canada | Hatchery | 0 | 45,009 | 1,738,916 | 1,783,925 |
|  | Wild | 0 | 9,083 | 329 | 9,412 |
|  | Adult | 0 | 6,046 | 0 | 6,046 |
|  | Total | 0 | 60,138 | 1,739,245 | 1,799,383 |
| Denmark | Hatchery | 72,900 | 0 | 0 | 72,900 |
|  | Wild | 0 | 0 | 0 | 0 |
|  | Adult | 0 | 0 | 0 | 0 |
|  | Total | 72,900 | 0 | 0 | 72,900 |
| Iceland | Hatchery | 127,162 | 0 | 0 | 127,162 |
|  | Wild | 2,516 | 0 | 0 | 2,516 |
|  | Adult | 0 | 563 | 0 | 563 |
|  | Total | 129,678 | 563 | 0 | 130,241 |
| Ireland | Hatchery | 289,029 | 0 | 0 | 289,029 |
|  | Wild | 939 | 0 | 0 | 939 |
|  | Adult | 0 | 0 | 0 | 0 |
|  | Total | 289,968 | 0 | 0 | 289,968 |
| Norway | Hatchery | 0 | 85,692 | 0 | 85,692 |
|  | Wild | 0 | 5,436 | 0 | 5,436 |
|  | Adult | 0 | 631 | 0 | 631 |
|  | Total | 0 | 91,759 | 0 | 91,759 |
| Russia | Hatchery | 0 | 3,000 | 417,750 | 420,750 |
|  | Wild | 0 | 40 | 190 | 230 |
|  | Adult | 0 | 1,809 | 0 | 1,809 |
|  | Total | 0 | 4,849 | 417,940 | 422,789 |
| Sweden | Hatchery | 0 | 4,928 | 39,517 | 44,445 |
|  | Wild | 0 | 0 | 0 | 0 |
|  | Adult | 0 | 0 | 0 | 0 |
|  | Total | 0 | 4,928 | 39,517 | 44,445 |
| UK (England \& | Hatchery | 100,537 | 5,061 | 65,858 | 171,456 |
| Wales) | Wild | 4,139 | 0 | 973 | 5,112 |
|  | Adult | 0 | 937 | 0 | 937 |
|  | Total | 104,676 | 5,998 | 66,831 | 177,505 |
| UK (N. Ireland) | Hatchery | 34,487 | 0 | 35,536 | 70,023 |
|  | Wild | 1,483 | 0 | 0 | 1,483 |
|  | Adult | 0 | 0 | 183 | 183 |
|  | Total | 35,970 | 0 | 35,719 | 71,689 |
| UK (Scotland) | Hatchery | 12,355 | 2,000 | 0 | 14,355 |
|  | Wild | 6,948 | 6,462 | 4,750 | 18,160 |
|  | Adult | 0 | 899 | 0 | 899 |
|  | Total | 19,303 | 9,361 | 4,750 | 33,414 |
| USA | Hatchery | 0 | 172,842 | 47,857 | 220,699 |
|  | Wild | 0 | 1,800 | 0 | 1,800 |
|  | Adult | 0 | 5,052 | 30 | 5,082 |
|  | Total | 0 | 179,694 | 47,887 | 227,581 |
| All Countries | Hatchery | 636,470 | 318,532 | 2,345,434 | 3,300,436 |
|  | Wild | 16,025 | 22,821 | 6,242 | 45,088 |
|  | Adult | 0 | 15,937 | 213 | 16,150 |
|  | Total | 652,495 | 357,290 | 2,351,889 | 3,361,674 |

Figure 2.1.1.1 Nominal catches of salmon in four North Atlantic regions 1960-2000.


Figure 2.1.3.1 Total reported catch, unreported catch (in NASCO Areas) and \% unreported catch of com bined catch 1986-2000


Figure 2.2.1.1 Worldwide farmed Atlantic salmon production, 1980 to 2000. Data for non-North Atlantic area do not include farmed salmon production in some countries ( notably Chile, which has a high production relative to other countries )


Figure 2.2.2.1 Production of ranched salmon in the North Atlantic, 1980 to 2000


Figure 2.3.1.1. Deterministic calculations of catch options for the fishery at West Greenland for low abundance and moderate abundance periods. Values in bold and in box are parameters with uncertainty.

| 2.6 Deterministic calculation of quota for a low abundance period |  |  |  |
| :---: | :---: | :---: | :---: |
| Step 1 | $\mathrm{SpR}=\mathrm{SpT}^{*}\left(\exp \left(11^{*} \mathrm{M}\right)\right)$ |  | $\mathrm{SpT}=2 \mathrm{SW}$ Conservation requirement for North America |
|  | $\mathrm{SpT}=$ | 152,548 |  |
| Step 2 | $\mathrm{SpR}=$ | 170,286 | $\mathrm{SpR}=$ Spawning Reserve for North America adjusted for 11 months of natural mortality between West Greenland and North America |
|  | $\mathrm{MAH}=\mathrm{PFA}-\mathrm{SpR}$ |  |  |
|  | PFA $=$ | 183,000 | PFA value at $50 \%$ probability |
| Step 3 | $\mathrm{MAH}=$ | 12,714 | MAH $=$ Maximum Allowable Harvest $=$ Number of surplus North American origin fish |
|  | NA SWW $=\mathrm{fNA} * \mathrm{MAH}$ |  | FNA $=$ fraction of NA surplus allocated to Greenland NA1SW = Number of North American surplus fish available for Greenland |
|  | FNA = | 0.4 |  |
|  | NA1SW = | 5,086 |  |
| Step 4 | E1SW $=($ NA1SW / PropNA) - NA1SW |  |  |
|  | PropNA = | 0.779 | PropNA = proportion NA salmon in the fishery |
|  | E1SW = | 1,443 | E1SW = number of European origin 1SW salmon expected in the fishery |
| Step 5 | Quota $(\mathrm{t})=($ NA1SW * WT1SWNA + E1SW * WT1SWE) * ACF / 1000 |  |  |
|  | WT1SWNA = | 2.666 kg | WT1SWNA $=$ weight $(\mathrm{kg})$ of 1SW NA origin salmon in the fishery |
|  | WT1SWE = | 2.832 kg | WT1SWE = weight (kg) of 1SW European origin salmon in the fishery |
|  | $\mathrm{ACF}=$ | 1.068 | $\mathrm{ACF}=$ age correction factor $(>=1)$ to account for fish other than 1SW of age |
|  | Quota $(\mathrm{t})=$ | 19 | Quota $=$ Allowable harvest ( t ) at West Greenland taking into account all the factors in steps 1 to 4 |


| 2.7 | Deterministic calculation of quota for a moderate abundance period |  |  |
| :---: | :---: | :---: | :---: |
| Step 1 | $\mathrm{SpR}=\mathrm{SpT}^{*}(\exp (11 * \mathrm{M})$ ) |  |  |
|  | $\mathrm{SpT}=$ | 152,548 | $\mathrm{SpT}=2 \mathrm{SW}$ Conservation requirement for North America |
|  | $\mathrm{SpR}=$ | 170,286 | $\mathrm{SpR}=$ Spawning Reserve for North America adjusted for 11 months of natural mortality between West Greenland and North America |
| Step 2 | $\mathrm{MAH}=\mathrm{PFA}-\mathrm{SpR}$ |  |  |
|  | PFA $=$ | 436,770 | PFA value at $50 \%$ probability |
|  | $\mathrm{MAH}=$ | 266,484 | MAH $=$ Maximum Allowable Harvest $=$ Number of surplus North American origin fish |
| Step 3 | NA1SW $=\mathrm{fNA} *$ MAH |  |  |
|  | FNA = | 0.4 | FNA $=$ fraction of NA surplus allocated to Greenland NA1SW = Number of North American surplus fish available for Greenland |
|  | NA1SW = | 106,594 |  |
| Step 4 | E1SW $=($ NA1SW $/$ PropNA $)-$ NA1SW |  |  |
|  | PropNA = | 0.59 | PropNA = proportion NA salmon in the fishery |
|  | E1SW = | 74,074 | E1SW = number of European origin 1SW salmon expected in the fishery |
| Step 5 | Quota $(\mathrm{t})=($ NA1SW * WT1SWNA + E1SW * WT1SWE) * ACF / 1000 |  |  |
|  | WT1SWNA = | 2.75 kg | WT1SWNA $=$ weight $(\mathrm{kg})$ of 1SW NA origin salmon in the fishery |
|  | WT1SWE = | 3.13 kg | WT1SWE = weight (kg) of 1SW European origin salmon in the fishery |
|  | $\mathrm{ACF}=$ | 1.068 | $\mathrm{ACF}=$ age correction factor $(>=1)$ to account for fish other than 1SW of age |
|  | Quota $(\mathrm{t})=$ | 561 | Quota $=$ Allowable harvest ( t ) at West Greenland taking into account all the factors in steps 1 to 4 |

Figure 2.3.1.2. Probability profiles for simultaneously achieving a given level of escapement relative to conservation in six stock areas of North America.


Figure 2.3.1.3. Probability profiles for the PFA forecast values for low abundance and moderate abundance periods.


Figure 2.3.1.4. Expected catch of 1SW salmon of North American origin at a catch option of 50 tons at West Greenland. The uncertainty in catch is quantified by incorporating the observed temporal variation in proportion of fish of North American origin, mean weights of 1SW salmon of North American and European origin, and the age correction factor for older age groups.


Figure 2.3.1.5. Theoretical risk analysis plots showing the risk-prone and risk-averse zones relative to the uncertainty of the stock assessment.


Figure 2.3.1.6. Examples of risk analysis profiles of catch options in West Greenland for low abundance and moderate abundance periods.


Figure 2.3.1.7. Relationship between small salmon (mostly 1SW salmon) in year i and large salmon (2SW salmon with an important component of multiple spawners) in year i+1, for 1985 to 2000.


Figure 2.3.1.8. Fishery, biological characteristics, and forecast data inputs to the risk analysis of the 2001 Miramichi homewater fishery.

## Assumptions of the fisheries risk analysis model



Figure 2.3.1.9. Risk analysis profiles for the 2001 homewater fisheries in the Miramichi River. The upper panel describes the egg loss from the harvest levels as a percentage of the total eggs in the predicted returns. The lower panel describes the risk to achieving the conservation requirements for different harvest levels

Egg Loss (\%)


Prob. of Meeting Conservation


Figure 2.4.3.1 Relationship between reported escapes and estimated escapees in Irish catch
(smolts are included with adults of the following year)



Figure 2.4.3.2. Geographical distribution of farmed salmon released at Bersagel. Grey and black dots are respectively river and sea recoveries.


Figure 2.4.3.3. Geographical distribution of farmed salmon released at Meløy. Red and black dots are respectively river and sea recoveries.


Figure 2.4.3.4 Ocean currents in the norteast Atlantic

Figure 2.4.4.1 Locations and numbers of post-smolts captured in surface trawl hauls during a special post-smolt survey tin mid- June 2000. Stars mark trawl-hauls without salmon. Black numbers indicate number of post-smolts and white numbers on black indicate number of older salmon in catch. Micro-tagged and Carlin- tagged fish were captured in the same catch of 170 fish, and a few micro-tagged fish were found on a neighbouring catch site (34). The approximate area where an international mackerel fishing fleet is operating during the summer months is shaded.


Figure 2.4.4.2 Sites of post-smolts captures 20 - 22 may 2000 during salmon lice investigations in the Sognefjorden and the Nordfjord. Numbers indicate post-smolts captured in Nordfjord, while stars indicate captures in the Sognefjord. Numbers of single catches are not known but the total Sognefjord catch was $>200$.


Figure 2.4.4.3 Average number of salmon lice per fish on seaward migrating post-smolts in trawl-captures in two southwest-Norwegian fjords in May 1998 - 2000.


Figure 2.4.5.1 The relationship between seasonal run-timing of returning adults and incidence of scales exhibiting summer checks. Data for 1SW salmon, 2 SW salmon exhibiting checks in their first year at sea and 2 SW salmon exhibiting checks in their second year at sea are shown separately, grouped by the year when the adults were taken by the net fishery (1997, white bars; 1998, stippled bars; 1999, black bars)




Figure 2.4.6.1. Predicted monthly mortality rate for Atlantic salmon relative to the sea age (days) as per the inverse weight relationship presented by Lorenzen (1996). River Bush data are from Doubleday et al. (1979).


Figure 2.4.6.2 Solutions for $S_{1}$ and $S_{2}$ for the hatchery returning Atlantic salmon from the Saint John River at Mactaquac (upper) and the wild smolts from de la Trinité River (lower).


Figure 2.4.6.3. Inter-river comparisons of annual survivals of the 1996 smolt cohort in year one (S1) (upper panel) and year two (S2) (lower panel). The Saint John smolts are hatchery age-1 smolts, the other two rivers are wild smolts.



Figure 2.4.6.4. Estimated monthly Z (instantaneous mortality rate) of salmon in their second year at sea. The points in grey (1992 and later smolt cohort) are likely a good representation of M since sea fisheries were declining or eliminated.


Figure 2.4.7.1. Mean annual air temperature trends for the Southwest and Northwest Miramichi monitoring stations.


Figure 2.4.7.2. Timing of the peak spring runoff event in the Miramichi River.


Figure 2.4.7.3. Mean annual fork length of juvenile Atlantic salmon from the Miramichi River, 1971 to 1999.




Figure 2.4.7.4. Association between mean annual size-at-age of age-1 and age-2 parr and the mean spring water temperature for the Miramichi River, 1971 to 1999.
 AREA

## 3.1 <br> Fishing at Faroes in 1999/2000

In the period 1991-98 inclusive the Faroese salmon quota was bought out. However, the Faroese Government continued sampling inside the 200 mile EEZ during most years (ICES 2000/ACFM:13). No buyout was arranged for 1999 and 2000. No fishing took place in 1999 and the commercial fishery resumed in 2000. In the 1999/2000 season approximately 8 t were caught by M/S "Túgvusteinur" during 2 commercial fishing trips between late January and early April 2000 (ICES 2000/ACFM:13).

### 3.2 Description of the 2000/2001 commercial fishery

No fishery for salmon was undertaken by Faroese fishermen during the 2000/2001 fishing season and, consequently, no biological information was available for this season.

### 3.3 Homewater Fisheries in the NEAC area

### 3.3.1 Significant events in NEAC homewater fisheries in 2000

In an effort to accelerate the phase out of mixed stock fisheries, Government funding ( $£ 0.75$ million) has been made available in UK (England and Wales) to facilitate buy-out initiatives. To be utilised, these 'pump-priming' sums have to be matched by equivalent funding from fishery and riparian interests.

UK (England and Wales), UK (Scotland) and Ireland experienced substantially higher than average rainfall in 2000; river flows were generally above average and provided good conditions for angling. However, flows were particularly high late in the year resulting in the loss of some fishing opportunities; these flows may also have impacted on spawning and recruitment (wash-out of redds).

In UK (Scotland), net fishermen agreed to suspend fishing for the first six weeks of the fishing season to protect the early running MSW (spring) salmon. This was a voluntary measure and was observed by the majority of fishermen.

In Ireland, two major draft net fisheries were suspended during 2000 as part of ongoing catchment management initiatives.

### 3.3.2 Gear

There were no reports of significant changes in the types of gear units used in the NEAC area countries in the year 2000.

### 3.3.3 Effort

The number of gear units licenced or authorised in several of the NEAC area countries is shown in Table 3.3.3.1. This provides a partial measure of effort but does not take into account other restrictions, for example, the close season. In addition, there is no indication from these data of actual number of licences utilised or the amount of time each licencee fished.

The change in effort for the last 5 years (1996-2000) for net fisheries is shown in Figure 3.3.3.1. In general there is a substantial decline in the net effort deployed in the NEAC area, the only exception being the driftnet fishery in Ireland.

The number of gear units licenced or authorised in the year 2000 declined in all but one of the countries where this information was available. In UK (England and Wales), UK (Scotland), Ireland, France and Norway the number of gear units licenced or authorised in 2000 was lower than in 1999. The only exception to this was the draftnet fishery in UK ( N -Ireland). Longer term trends were also consistent among these countries. In all cases, the number of gear units licenced or authorised were at the same level or lower than the previous 5- and 10-year averages with exception for the driftnet fishery in Ireland where an increase occurred compared to the previous 5-and 10-year averages.

Figure 3.3.3.2 show the change in effort over the last 5 years (1996-2000) for rod fisheries in the NEAC area countries where effort is reported. The rod fishery effort has decreased in UK (England and Wales) but increased in Ireland, Finland and France.

There was an increase in the number of rod licences issued in Ireland compared to 1999 and also compared to the previous 5-year average. In Finland, rod effort in the River Teno recreational fishery increased substantially (41\%) compared to 1999 and to the previous 5- and 10-year averages. In River Naatamo the number of fishermen decreased compared to the previous year and to the previous 5- and 10-year averages. A similar decline in effort was evident in the total declared number of rod days fished in UK (England \& Wales) with the 2000 value down compared to previous indices.

### 3.3.4 Catches

NEAC area catches are presented in Table 3.3.4.1. The total catch in the NEAC area was 2633 tonnes, up $27 \%$ on the 1999 catch, and representing $94 \%$ of the total North Atlantic nominal catch in 2000. Both homewater and total reported catches in NEAC area showed increases compared to 1999 ( 27 and $25 \%$ respectively). This increase in total catches arises from substantial increases of the nominal catch in a few countries (Norway, Finland, and Sweden), while others showed only slight increases or even substantial decreases (UK (Scotland) and Iceland). The nominal catches for individual countries can be found in Table 2.1.1.1. Figure 3.3.4.1 shows the percentage change in the 1999 NEAC homewater catches relative to the previous 5-year (1995-99) and 10-year (1990-99) means.

As can be seen in Figure 3.3.4.1, four countries (UK (England \& Wales), Sweden Norway and Finland) showed a substantial ( $20-91 \%$ ) increases in catches relative to the 5 -year averages, while two of them (Norway and Finland) also showed large ( $40-65 \%$ ) increases in relation to the 10 -year average. These changes are believed to arise from higher marine survival rates rather than an increase in effort (cf. Section 3.4). The catches in UK (Scotland) and Iceland decreased substantially ( $>35 \%$ ) in relation to both long-term averages, while Russian and French catch decreased only relative to the 5 -year average. For the remaining NEAC countries reporting catches greater than 10 tonnes the changes were minor.

### 3.3.5 Catch per unit effort (CPUE)

CPUE data for the NEAC area are presented in Tables 3.3.5.1-3.3.5.5. The CPUE data for rod fisheries has been collected by relating the catch to rod days or angler season, and that of net fisheries was calculated as catch per licenceday, trap month or crew month. Grouping of data for the trend analysis was based on the units in which CPUE was presented (rod fisheries) and on national/regional distribution (rod and net fisheries) of fisheries considered.

It is assumed that the CPUE of net fisheries is a more stable indicator of the general status of salmon stocks than rod CPUE; the latter may be more affected by varying local factors, e.g. weather conditions, management measures and angler experience. However, both may also be affected by many measures taken to reduce fishing effort, for example, changes in regulations affecting gear.

In Finland, CPUE in the rivers Teno and Naatamo showed increases over the previous year and, in the case of the Teno, an increase over the previous 5 year average. In France, there was a small increase over the previous year but a decrease on the previous 5 year average. The river Bush (UK (NI)) showed a decrease compared to both recent indices (Table 3.3.5.1). Information from eight rivers in Russia showed, in most cases, increases over both 1999 and previous 5 year averages (Table 3.3.5.2). Results of the route regression analysis showed no general trends in rod CPUE over the last 10 years (Table 3.3.5.6).

In UK (England and Wales) CPUE for the net fishery increased in the North East and North West regions compared to 1999 and the previous 5 -year averages. In contrast, the Wales, South West and Midland regions showed a decrease compared to recent indices (Table 3.3.5.3). Route regression analysis showed a significant upward trend for the last 10years in these fisheries. The CPUE for the Scottish net fisheries showed an increase over 1999 but values remained lower than the previous 5 -year mean (Table 3.3.5.4). Route regression analysis for the Scottish net fishery revealed a significant downward trend for the last 10 -years (Table 3.3.5.6).

CPUE for the marine fishery in Norway has increased for the past three years for bagnets and bendnets and was consistent across all size groups (Table 3.3.5.5). The CPUE was highest for small fish ( $<3 \mathrm{~kg}$ ).

CPUE is a measure that can be influenced by various factors. Water level and weather condition can have effects on CPUE as well as operation intensity or number of gear units in an area. If large changes occur for one or more factors a
common pattern may not be evident over larger areas. It is, however, expected that for a relatively stable effort CPUE can reflect changes in the status of stocks and stock size. This can be seen in the increase in CPUE for the Norwegian marine fishery that is also reflected in increased catch (Section 3.3.4) as well as the calculated PFA values (Section 3.6.).

### 3.3.6 Age composition of catches

The percentage of 1 SW salmon in catches is presented in Table 3.3.6.1 and Figure 3.3.6.1 for those countries where a time series of data exist. The proportion of 1SW fish in the 2000 catches is presented as a percentage of the 1995-99 and 1990-1999 averages. In comparison with 1999 the differences between countries in Northern and Southern Europe were less apparent in 2000, with southern European countries occurring at both ends of the scale (Figure 3.3.6.1). While the proportion of 1 SW fish in the 2000 catch remained lower than both the 5 and the 10 year average in UK (Scotland), it had changed from lower to higher than the long-term indices in France and UK (England and Wales). Several NEAC countries also report nominal catches partitioned according to sea-age category (Table 2.1.1.3.). These data suggest that for France the change in the age composition of the 2000 catch was driven by the combined effect of an increase in 1 SW catches and a reduction in the number of MSW fish taken.

There was no common trend in the age composition of the 2000 catches of northern European countries (Finland, Iceland, Norway, Sweden, and Russia), as two had gone down and three had remained higher than the long-term averages.

### 3.3.7 Farmed and ranched salmon in catches

The contribution of wild, farm-origin and ranched salmon to national catches in the NEAC Area 1991-2000, is shown in Table 3.3.7.1. Although showing a slight decrease in 2000, farmed salmon continued to account for a relatively large proportion ( $20 \%$ ) of the nominal catch in Norway. Farmed salmon was estimated to constitute less than $2 \%$ in all other NEAC countries.

Table 3.3.7.2 gives estimates of the incidence of farmed salmon in Norwegian coastal and fjord fisheries. In 2000, farmed fish accounted for $31 \%$ of the catch of coastal fisheries, a decline compared to the previous 5 years estimates. In 2000, the proportion of farmed salmon in Norwegian rod catches was similar to the preceding year, whereas the incidence of farmed fish in brood stock samples was estimated to be $8 \%$, a reduction of almost $50 \%$ (Table 3.3.7.3). This decrease in the proportion of farmed salmon in the catches is probably due to a rise in the Norwegian catches of wild fish rather than to a true decrease in escapees, as the number of farmed salmon captured was the highest in the whole time series.

In the River Teno (Finland and Norway), the incidence of farmed salmon during the fishing season in June-August has been low, varying between $0.04 \%$ and $0.4 \%$ over the period 1987 to 2000 . However, occasional samples taken in 1990 - 91 and 1997 after the fishing season (September-October) have included 0 to $50 \%$ farmed fish (Table 3.3.7.4), indicating that the proportion of fish farm escapees in the spawning stock may be higher than shown in the in-season samples. This is similar to data presented from Norway (Table 3.3.7.3).

Catches of salmon in coastal fisheries in both UK (Northern Ireland) and Ireland are examined for escaped farmed salmon (Table 3.3.7.5). Data for both countries are presented together as they constitute a continuous part of the species' geographic range. Escaped farmed fish have been detected every year; the combined frequency being less than $1 \%$ in most years. The 2000 figures for two specific areas (UK Northern Ireland and the Galway Limerick area in Ireland), however, show slightly elevated levels ( $3.1 \%$ and $2.2 \%$ respectively) compared to previous years.

In UK (Northern Ireland), only $0.3 \%$ of the total salmon run trapped in the River Bush comprised farmed salmon, continuing the trend for a low incidence of farmed fish being detected in freshwater (Table 3.3.7.6).

A catch sampling programme in UK (Scotland) from 1981 to the present indicates that the incidence of farmed salmon in catches of fisheries around the country continues to decrease from their highest recorded levels around 1993 (Table 3.3.7.7).

The incidence of farmed salmon in catches is examined in relation to farmed production in Section 2.4.3.

Some new information on tag recoveries was made available to the Working Group. From 1996 to 1999 a total of 409,762 smolts, mainly hatchery reared were tagged and released in Norway. A total of 3,811 adult recoveries were reported from Norway and 19 from other countries $(0.5 \%$ of the total number of salmon recovered). This is consistent with previous observations that very few Norwegian salmon are intercepted in other countries.

### 3.3.9 Exploitation rates in homewater fisheries

The analyses of exploitation rates were greatly expanded in 2000 as data from 15 new rivers were added. Exploitation rates for 25 wild, 4 mixed (wild and hatchery) and 12 hatchery stocks are shown in Table 3.3.9.1. Route regression analysis shows no trend relative to any of the long-term means for the 1SW stock components from the rivers reported. The analysis detected a significant downward trend in exploitation for the past 10 -year periods for the 2 SW-component of four of the south- European rivers, two southwest-Norwegian and one Icelandic and the Swedish river included in the analysis. Exploitation rates for these rivers showed no trend for the 2 SW component on a 5 year period. As in 1999, exploitation rates of all ages of salmon in the Russian rivers draining to the White Sea Basin showed a significant downward trend both for the 5 and the 10 year period. Among the rivers draining to the Barents Sea, only the exploitation rates for the 10 -year period showed a downward trend while the 5 -year period showed no trend (Table 3.3.5.4).

### 3.3.10 Summary of homewater fisheries in the NEAC area

In the NEAC area there has been a general reduction in catches since the 1980s. This reflects a decline in fishing effort, both as a consequence of management measures and the reduced value of commercially caught salmon, as well as a reduction in the size of stocks. However, the overall nominal catch in the NEAC area in $2000(2,633 \mathrm{t})$ was substantially higher than that in 1999 (2074t). Catches in some northern European countries were particularly high and well above the recent five and ten-year averages; catches in most other NEAC countries were close to or below the long-term averages.

While there have been no changes in the types of commercial fishing gear used, the number of licensed gear units has, in most cases, continued to fall. Most fisheries for which data are available record a reduction of over $40 \%$ in gear units operated over the last 10 years. There are no such consistent trends for the rod fishing effort in NEAC countries over this period. Further initiatives to reduce fishing effort were introduced in several countries.

CPUE data for the net and rod fisheries show differences between countries but no large scale geographic patterns emerge. The Working Group noted that reduction in the number of fisheries operating can benefit those fisheries still in operation and that the lack of consistent trends in CPUE may reflect the imprecise nature of these indices.

Exploitation rates showed no trends relative to long-term averages for 1SW stock components in the NEAC area, although significant downward trends were detected for the 2SW component of some fisheries. There appeared to be no uniform pattern across NEAC countries.

No common trends were noted in the sea age composition of the 2000 catches in the NEAC areas. Differences in the age composition between countries in Northern and Southern Europe noted in 1999 were less apparent in 2000.

In general, the incidence of farmed salmon in NEAC homewater fisheries remained at low levels ( $<2 \%$ ) and similar to recent years, despite the continued increase in the salmon farming industry. The proportion of farmed salmon (20\%) in the nominal catch for Norway did not increase, but this was a result of the significantly increased catch of wild fish. The number of farmed fish in the Norwegian catch was the highest recorded in the time series.

### 3.4 Status of Stocks in the NEAC Area

### 3.4.1 Attainment of conservation levels

Attainment of conservation limits was examined for 13 rivers in the NEAC area for which spawning escapement data (egg deposition) and river-specific conservation limits were available. The analyses included five rivers from Russia, four from UK (England \& Wales), and one each from UK (Northern Ireland), UK (Scotland), Ireland and France, where sufficient data were available for the previous 10 years (Figure 3.4.1.1). However, this set of rivers does not sufficiently represent the entire NEAC area as, for instance, no data are available from Norway. In contrast, from 45 rivers in UK
(England \& Wales) where CL attainment data were available, five rivers were selected for the analyses to represent different regions.

In accordance with the analysis carried out last year, four categories of rivers were distinguished (Fig. 3.4.1.2):

- Type A: four rivers in which egg deposition was mostly below CL (means of the river CL attainment rates of $45 \%$ to $80 \%$ );
- Type B: four rivers in which egg deposition fluctuated around CL (means of the river CL attainment rates of $84 \%$ to $119 \%$ );
- Type C: three rivers in which egg deposition was mostly above CL (means of the river CL attainment rates of $153 \%$. to $180 \%$ );
- Type D: two rivers in which egg deposition was well above CL (means of the river CL attainment rates of $250 \%$ and $281 \%$ ). These rivers never fell below CL during the last 10 years.

Information about the conservation limit attainment for the 13 rivers and one river with a shorter time series available (Coquet, UK England) is given in Table 3.4.1.1. and Fig. 3.4.1.1. Eight rivers ( $62 \%$ ) out of the 13 rivers where data were available for the year 2000 showed a decline from 1999 to 2000 in the CL attainment (Table 3.4.1.1, Fig. 3.4.1.1.), and in five cases the reduction varied between $30 \%$ and $84 \%$. Rivers with the largest decline belong to the two first categories defined above (Fig. 3.4.1.2). In contrast, three rivers showed a marked increase in their CL attainment (27$57 \%$ ) all belonging to different categories (Fig. 3.4.1.2). Seven rivers (54\%) had an egg deposition above their CL in 2000, which is more than in 1999 ( $44 \%$ ) but less than in $1998(69 \%)$.

The slight increase in CL attainment from that of the previous year was in accordance with the increase in catches (Section 3.3.4), although the CL attainment information available is not representative of the entire NEAC area.

### 3.4.2 Measures of juvenile abundance

Smolt counts or estimates of juvenile abundance were made available to the Working Group for 14 rivers (Table 3.4.2.1). Estimates of juvenile salmon ( $0+$ ) abundance in the rivers Bush (UK N.Ireland) and Nivelle (France) were clearly higher than the 5 -year mean being the highest (Nivelle) or second highest (Bush) for the last 10 years. In the River Teno and its tributaries (Finland), juvenile salmon (fry \& parr) abundance was mostly at or below the previous 5year mean and below the figures for the previous year (Table 3.4.2.1).

About half of the smolt counts in 2000 were higher than those of the previous year and the 5 -year means (Table 3.4.2.1). There was a significant downward trend during the last 10 years for the smolt counts in all rivers together ( $\mathrm{p}=0.981$ ) and separately for the northern rivers (Norway, Sweden; $\mathrm{p}=0.984$ ), whereas no trends were detected during the last 5 years for these groups of rivers. No trends were detected for southern rivers (UK N-Ireland, Ireland, France) for the 5- or 10- year periods (Table 3.4.2.2).

In general, the declining smolt counts in northern rivers appear to reflect the previous decline in adult returns and variable attainment of the CL in this area. However, the downward trend in returns and PFA estimates of southern rivers does not seem to be reflected in smolt counts of the southern rivers.

### 3.4.3 Measures of adult returns back to the rivers

Estimates or total counts of adult salmon returning into the rivers were available for 36 rivers in the NEAC area (Table 3.4.3.1). To examine trends in different geographical groups of rivers, the rivers were divided into northern (Nordic countries and Russia) and southern group (UK, Ireland and France). Information from 32 rivers was available to examine the change between 1999-2000 and in relation to the previous 5-year mean.

In the northern group, eight out of 14 rivers ( $57 \%$ ) showed an increase from the previous year but only six rivers ( $43 \%$ ) exceeded their 5 -year mean. A significant downward trend was detected for the previous 10 -year period in the northern group ( $\mathrm{p}=0.99$ ), whereas no trend was detected for the previous 5 -year period $(\mathrm{p}=0.32)$.

In the southern group, more than half ( $56 \%$ ) of the adult salmon counts were lower than those in 1999, and two thirds ( $66 \%$ ) were below their 5 -year mean (Table 3.4.3.1). Due to technical limitations of the route regression analysis, it was necessary to split the southern rivers into two groups: British Isles ( $\mathrm{n}=17$ ) and France ( $\mathrm{n}=4$ ). A significant downward trend was detected for the previous 10 -year period in both groups (British Isles: $p=0.97$, France: $p=0.98$ ). The rivers in

France showed a downward trend for the previous 5 -year period $(\mathrm{p}=0.97)$ whereas no such trend was detected for the rivers in UK and Ireland ( $\mathrm{p}=0.74$ ).

### 3.4.4 Survival indices

Estimates of marine survival for wild smolts from 5 stocks returning to homewaters (i.e. before homewater exploitation) for the 1999 smolt year class are presented in Table 3.4.4.1. For the Nivelle River (France), indices of survival are also provided for autumn age $-0^{+}$parr. This provides an approximation of marine survival as more than $80 \%$ of juveniles migrate after only one year in freshwater. In most rivers marine survival for the 1999 smolt year class was below the 5and the 10 -year mean for both 1 SW and 2 SW fish. Route regression analysis showed significant downward trend in marine survival for 1 SW fish for the last 5- and 10 -year period ( $\mathrm{p}=0.95$ and 0.99 ), while no trend was detected for 2SW fish.

Marine survival for 13 hatchery stocks are given in Table 3.4.4.2. For the past 10 -year period, route regression analysis showed a significant downward trend for survival to homewaters for 1 SW ( $\mathrm{p}=0.983$ ) and 2 SW fish ( $\mathrm{p}=0.998$ ), but no trend was detected for the past 5-year period for either 1SW or 2SW fish. However, return rates of hatchery released fish may not always be a reliable indicator of marine survival of wild fish because of differences in release conditions.

These results are consistent with the information on adult salmon counts (Section 3.4.3), and suggest that returns are strongly influenced by factors in the marine environment.

### 3.4.5 Summary of the status of stocks in the NEAC area

Analysis of attainment of conservation limits (CL) in 2000 showed that the proportion of rivers with an egg deposition above their CL was higher than in 1999 but less than in 1998. However, a majority of rivers showed a decline in their level of attainment in 2000 compared to the previous year and in most cases the decline was substantial ( $30-80 \%$ ). This indicates that the recovery of salmon stocks observed in 1998, from a period of low attainment (1994-1997), has not appeared to continue. Although some areas were not represented in the data (e.g. Norway), the Working Group had no reason to assume that the indices were not representative of stocks in general and noted the analysis broadly corresponded to the results of the PFA analysis in Section 3.4.6.

Measures of smolt production indicated that about half of the rivers showed higher smolt output in 2000 than in 1999. Route regression analysis revealed a significant downward trend over the last 10 years for northern rivers, but no trend was detected for southern rivers. The declining smolt counts in northern rivers appear to reflect the previous decline in adult returns and variable attainment of the CL in this area. However, the downward trend in returns and PFA estimates of southern rivers does not seem to be reflected in smolt counts of the southern rivers.

Measures of adult returns back to the rivers showed that of the rivers examined in 2000, more than half showed increased counts. Both southern and northern rivers showed a significant decline over the last 10 years, whereas no clear trend was detected for the last 5 -year period.

For most rivers where information is available, marine survival indices were below both the previous 5 -and 10 -year means. Route regression analysis showed significant downward trend in marine survival for 1SW fish for the last 5-and 10 -year period, while no trend was detected for 2 SW fish. A similar analysis showed a downward trend in marine survival for 1 SW and 2 SW hatchery fish over the last 10 years but no decline over the past 5 year period. These results are consistent with the information on adult salmon counts and suggest that returns are strongly influenced by factors in the marine environment.

In summary, the monitored rivers analysed in this section would suggest that the status of salmon stocks in the NEAC area is, in general poor. This broadly agrees with the results of the PFA-lagged spawner analysis that is based on national catch statistics and and presented in Section 3.6.

### 3.5 Evaluation of the effects on stocks and homewater fisheries of significant management measures introduced since 1991.

### 3.5.1 Evaluation of the Effects of the Suspension of Commercial Fishing Activity at Faroes

Between 1991 and 1998 the Faroese fishermen agreed to suspend commercial fishing for the salmon quota set by NASCO, in exchange for compensation payments. The number of fish spared as a result of this period of suspension is
the catch that would have been taken if the fishery had operated, minus the catch in the research fishery which operated in most years. No buyout was arranged for 1999 or 2000. Although no fishing took place in 1999, a single vessel carried out commercial fishing in 2000, catching approximately 8 t . As for last year (ICES 2000/ ACFM:13), analysis was based on the assumption that full quota would have been taken, had full scale commercial fishing taken place. Thus, the maximum catch that would have been taken in 1999/2000 would have been 300 t (Table 3.2.1.1). For the 1999/2000 analysis therefore the fish spared totalled 292t (300t-8t).

Although commercial fishing was carried out in 1999/2000, no new data on the discard rates, the age composition of the catch, or for the proportion of farm escapees in catches were available. Hence, the same values were used as for the 1998/99 assessment. The assessment is shown in Table 3.5.1. This suggests that if the full quota had been bought out, between 3,000 and 21,000 additional 1SW salmon and between 70,000 and 138,000 additional MSW salmon would have returned to homewaters each year from 1992 to 2000. For the 1999/2000 season, the numbers of fish believed saved were 15,332 1SW and $87,726 \mathrm{MSW}$, respectively. In addition, between 27,000-55,000 escaped farmed fish each season would have been saved from capture in the Faroese fishery. However, data from tagging experiments suggest that almost all survivors would return to Norway (Hansen and Jacobsen, 1997), provided they behaved in a similar manner to wild fish. The analysis carried out suggests that, for the 1999/2000 season, an estimated 24,000 escapedfarmed fish may have been saved.

Estimates (means of 1000 simulations) of the total numbers of 1SW and MSW salmon returning to homewaters (i.e., Pre Fishery Abundance estimates) in the NEAC area and to countries of northern and southern Europe are provided in Tables 3.6.4.1 and 3.6.4.2. The calculated additional returns represent between $7 \%$ and $15 \%$ of MSW fish and up to $1 \%$ of 1SW fish returning to homewaters between 1992 and 2000 (Table 3.5.1). However, data from adult tagging studies (Hansen and Jacobsen 1997), indicate that the majority (about 65\%) of MSW salmon caught in the Faroes fishery would return to Scandinavian countries, Finland and Russia. If this were the case, they might have represented from $10 \%$ to $20 \%$ of MSW returns and up to $2 \%$ of 1SW returns to northern European homewaters in the same period (Table 3.5.1). If stocks and fisheries had remained stable, total catches would have been expected to increase by approximately the same proportions in respective areas.

Although the assessment of changes in catch levels prior to and after suspension of the fishery at Faroes were influenced by many other factors (eg. changes in exploitation and effort) could have affected abundance, the benefits of not fishing should be highlighted. The additional returns will still be expected to contribute to catches and spawning stocks, even if any expected increase has been masked by other factors such as changes in marine survival or management measures in homewaters, such as those outlined in Section 3.5.2, below.

### 3.5.2 Evaluation of the effects of management measures introduced in homewaters since 1991.

The Working Group noted significant reductions in the number of gear units deployed in most countries in the NEAC area (Table 3.5.2). Additional measures have been taken is some countries.

In Ireland new management and conservation legislation was brought into force in 1997 which was aimed at reducing effort and exploitation in the fishery and to facilitate enforcement. These regulations were enforced during the 19972000 seasons. In order to show whether there has been a change in the catch subsequent to the introduction of measures in 1997, the data were analysed using a Non-Parametric Random Ratio (NPR) test (Rago 1993, ICES 2000/ACFM:13). This test compares the mean catch in the period following the introduction of new regulatory measures in 1996 (i.e. 1997-2000) with the mean catch reported in the preceding 7 years to 1990. The results of this test indicate that drift net catches in the most recent 4 years were significantly lower than the preceding 7 years ( $\mathrm{p}<0.01$ ) in all regions. Similarly, draft net catches (excluding the North Western Region where the Moy River draft net was suspended in 1994) were also significantly lower in the most recent 4 years ( $\mathrm{p}<0.01$ ) in all regions. A Non Parametric Random (NPR) Ratio Test (Rago, 1993, Anon 2000) was applied to the exploitation data for 10 Irish stocks to examine the effects of the management measures introduced in 1997. The results of this test indicated that the exploitation rates in the period from 1997 to 2000 were also significantly reduced compared to the previous 10 year period ( $\mathrm{p}<0.01$ ). It is concluded therefore that the measures introduced in 1997 contributed to a reduction in both the overall catch and the exploitation rate on Irish stocks.

In UK (England and Wales), the North East coast fishery is the largest net fishery in UK (England and Wales) and has taken $68 \%$, on average, of the national declared net catch over the period 1970-92. A phase out of this fishery was introduced in 1993 and the number of licences issued has subsequently fallen by $50 \%$, from 142 in 1992 to 71 in 2000. The exploitation rate in 1992 was estimated to be in the region of $50 \%$. Assuming the remaining fishermen are representative and that there have been no major changes in the fishery, the average exploitation rate (1996-2000) would have fallen to around $32 \%$ (i.e. a $36 \%$ reduction). This is in close agreement with the reduction in the average
drift net catch (1996-2000), which has fallen by $32 \%$ compared with the 5 years (1988-92) prior to the start of the phase out. A number of other smaller coastal mixed stock fisheries have also been phased out since 1991.

National measures introduced in UK (England and Wales) in 1999 to protect spring salmon are estimated to have saved around 3,700 salmon from capture by net fisheries in 2000 before June 1 (based on the catch and the average proportion of fish taken in this period in the 5 years prior to the measures being introduced) and 1,400 by rod fisheries (based on a similar proportion, but adjusted for catch and release).

The Working Group noted that a large number of other measures had been introduced. For example, in UK (England and Wales) the total number of licences issued has been reduced by $46 \%$ since 1991, but the introduction of additional controls (e.g. increased close periods) has reduced the total allowable fishing effort by $60 \%$. In UK (Scotland), a voluntary cessation of net effort for the first six weeks of the fishing season was agreed by the majority of fishermen and was introduced in 2000. In the rod fishery, there has been an increase in the practice of catch and release since 1994 as previously reported (Table 2.1.2.1). These recent initiatives, and the continuing decline in overall net effort, are likely to reduce the impact on NEAC stocks. The Working Group expected these changes to reduce homewater exploitation rates.

### 3.6 Expected abundance of Salmon in the North East Atlantic

### 3.6.1 Previous development of a NEAC - PFA model

The Working Group has previously developed a model to estimate the pre-fishery abundance (PFA) of salmon from countries in the NEAC area. (PFA in the NEAC area is defined as the number of 1SW recruits on January $1^{\text {st }}$ in the first sea winter). The method employs a basic run-reconstruction approach similar to that described by Rago et al. (1993) and Potter and Dunkley (1993). The model estimates the PFA from the catch in numbers of 1SW and MSW salmon in each country. These are raised to take account of minimum and maximum estimates of non-reported catches and exploitation rates of these two sea-age groups. Finally these values are raised to take account of the natural mortality between January $1^{\text {st }}$ in the first sea winter and the mid-point of the respective national fisheries. A Monte Carlo simulation (1000 runs) (using 'Crystal Ball' in Excel; Decisioneering, 1996) is used to estimate confidence limits on the PFA values. Full details of the model are provided by Potter et al. (1998).

### 3.6.2 Improvements to the NEAC-PFA model

No changes were made to the structure of the PFA model used in 2001. In addition to adding catch data and other parameter values for 2000, national representatives considered methods for improving the precision of parameter values used in the model, in particular by splitting the national data set into regions. The following changes were made to the input data sets:

| Country: | Changes to data inputs: |
| :--- | :--- |
| Finland | No changes; national data set applies to only 2 rivers. |
| France | No changes made to data inputs; a new model for estimating numbers of returns is <br> being developed. |
| Iceland | National data set split into south-western and north-eastern regions for 1971-2000 |
| Ireland | Seven regional data sets were evaluated but their use in the assessment was <br> considered inappropriate at this stage; the full data set was reviewed and <br> exploitation and non-reporting rates were modified. |
| Norway | National data set split into 3 regions (South , middle and North) for period 1983- <br> 2000. |
| Russia | Data will be split into 3 regions for 2001; methods for handling 'fall salmon' <br> (salmon that start to enter freshwater more than one year before spawning) will <br> also be reviewed. |
| Sweden | Data will be fully re-evaluated for 2001 |
| UK(England \& Wales) | Data set fully reviewed and corrected; a new model for combining counter and <br> catches data to estimate total returns is being developed. |
| UK(Northern Ireland) | National data set split into 2 regions (Foyle and Fisheries Conservation Board) for <br> 1971-2000 |
| UK(Scotland) | National data set split into eastern and western regions for 1971-2000 |

The national and regional input data sets listed above for countries in the NEAC area and for the Faroes and West Greenland fisheries are shown in Appendix 7. The maximum and minimum values denote the limits of the uniform distributions used in the Monte Carlo Simulations.

Tables 3.6.2.1 to 3.6.2.6 summarise the outputs from the simulation, giving the mean estimates for each NEAC country of the numbers (plus variances/standard deviations) of:

- returns (1SW and MSW) (Table 3.6.2.1 and 3.6.2.2),
- recruits (PFA) (maturing and non-maturing 1SW and total 1SW) (Table 3.6.2.3 and 3.6.2.4),
- $\quad$ spawners (1SW and MSW) (Table 3.6.2.5 and 3.6.2.6).

Trends in these outputs are discussed in Section 3.6.4.

### 3.6.3 Grouping of national stocks

The Working Group has previously considered the most appropriate stock groupings for the provision of catch options or alternative management advice. Stock groupings may be proposed for two purposes: first, to provide advice on the status of all the stocks in the NEAC area in a way that can conveniently be used by NASCO managers; and second to provide catch advice for the stocks contributing to specific fisheries. Different stock groupings may be appropriate to fulfil these objectives.

No new information was presented to the Working Group on methods to group stocks in the NEAC assessment, although it was noted that this is one of the topics being considered by a current EU Concerted Action programme ("SALMODEL"). The Working Group has therefore continued to use the following groups of countries to present the PFA data:

| Southern European countries: | Northern European countries: |
| :--- | :--- |
| Ireland | Finland |
| France | Norway |
| UK(England \& Wales) | Russia |
| UK(Northern Ireland) | Sweden |
| UK(Scotland) | Iceland |

### 3.6.4 Trends in the PFA for NEAC stocks

Tables 3.6.2.1 to 3.6.2.6 show combined results from the PFA assessment for the Northern and Southern European groups and the whole NEAC area. The PFA of maturing and non-maturing 1SW salmon and the numbers of 1SW and MSW spawners for these areas are shown in Figures 3.6.4.1 to 3.6.4.6.

The $95 \%$ confidence limits (dotted lines for PFA and vertical bars for the spawning escapement) shown in Figures 3.6.4.1 to 3.6.4.3 indicate the level of uncertainty in this assessment procedure. In more recent years this level of uncertainty has decreased. However, the Working Group recognised that the model provided an interpretation of our current understanding of national fisheries and stocks based upon simple parameters. Errors or inconsistencies in the output must largely reflect errors in our best estimates of these parameters. Furthermore, there are risks that progressive errors could occur if, for example, the rate that exploitation has been reduced over a period of years is underestimated. The results therefore need to be treated with caution.

Figure 3.6.4.1 shows that there has been a general decline in recruitment of 1 SW and MSW salmon in the whole NEAC area over the past 30 years, and both age groups have fallen to their lowest levels in the three years. Numbers of 1SW and MSW spawners have also declined (Figure 3.6.4.2) over the past 30 years, although the decline has been less severe, indicating that reductions in exploitation have, to some extent, compensated for the decline in stocks.

Figure 3.6.4.3 shows that recruitment of maturing 1SW salmon (potential grilse) in Northern Europe was generally high (around one million) in the 1970s and 1980s, although the numbers have fluctuated quite widely, but there was a steady decline in these stocks from the mid 1980s to the mid 1990s. In the past four years there has been an upturn in the recruitment, with stocks in 2000 returning to the levels observed in the early 1990s. The number of 1SW spawners was
low in the 1970s, increased through the 1980s but declined again in the 1990s (Figure 3.6.4.4). However, escapement in 2000 appears to have been good.

Numbers of non-maturing 1SW recruits (potential MSW returns) for Northern Europe are also estimated to have fluctuated around one million between 1970 and 1985, but subsequently fell to about half this level in the late 1990s; there has been a slight upturn in the past three years. The numbers of MSW spawners, however, show no trend over the time series although numbers appear to have been good in 2000. It therefore appears that the decline in recruitment has been balanced by the reductions in exploitation both in homewater fisheries and at Faroes. These trends in recruitment for the Northern European stocks are broadly consistent with the limited data available on the marine survival of monitored stocks in Norway and Iceland (Section 3.4).

In the Southern European stock complex (Figure 3.6.4.5), the numbers of maturing 1SW recruits are estimated to have fallen substantially since the 1970s. Recruitment was at its lowest during the 1990s and there was a further drop in the estimated recruitment in 1999 with value in 1999 and 2000 being the lowest in the time series. This pattern is consistent with the data obtained from a number of monitored stocks. Survival of wild smolts to return as 1SW fish fell to very low levels on the four monitored rivers in the Southern European area for which data were available (Corrib and Burrishoole (Ireland), Bush (UK(Northern Ireland)), Nivelle (France)) (Section 3.4). This suggests that the marked reduction in 1SW returns in 1999 is likely to have been due in large part to a widespread decline in marine survival. Reductions have also been observed in freshwater production and marine survival could be affected by factors operating in freshwater.

The PFA estimates suggest that the number of non-maturing 1SW recruits in Southern Europe has declined fairly steadily over the past 30 years (Figure 3.4.6.5); these stocks have also reached their lowest levels in the time series in 1999 and 2000. This is broadly consistent with the general pattern of decline in marine survival of 2SW returns in most monitored stocks in the area. In more recent years, reductions in exploitation do not appear to have kept pace with the stock declines and the spawning escapement has thus also fallen over the period (Figure 3.6.4.6).

### 3.6.5 Forecasting the PFA for NEAC stocks

In order to provide numerical catch advice, PFA values must be forecast one or two years in advance. For example, the PFA of non-maturing 1SW recruits must be predicted for 2000 if we are to provide catch advice for the West Greenland fishery in 2001, for the Faroes fishery (MSW stock) in the 2001/02 fishing season and for homewater MSW fisheries in 2002. Because the latest estimate of non-maturing 1SW recruits is for 1999, the PFA must be forecast two years ahead, as is currently practised for the North American assessment. For maturing 1SW stocks, a single year's projection is sufficient.

The Working Group currently has no method for predicting future PFA levels, other than by extrapolation from the estimated time series which due to the uncertainties is not considered to be appropriate. Co-ordination of work on the PFA assessments and predictive models for NEAC stocks is being considered as part of an EU Concerted Action programme and is expected to be one of the more important deliverables resulting from this Concerted Action. .

### 3.6.6 Evaluation of effects of farmed salmon on the model

NASCO has asked ICES to evaluate the potential biases in the catch advice resulting from the inclusion of farmed escapees in the assessment models. The NEAC PFA model has previously only taken account of the presence of farm escapees in the Faroese catches. The estimated proportion of wild fish in the Faroese catches is shown in Appendix 8h; values for the years 1984 to 1997 were based on scale sampling (Hansen, et al. 1999), while values have been assumed or estimated for other years on the basis of the level of farm production and catches. The recruits back-calculated from the Faroese catches are thus divided into wild and farmed components; the wild fish are re-allocated to the national estimates of 1SW (maturing and non-maturing) recruits according to proportions estimated from various tagging studies; farm escapees in the Faroese catches are ignored in the assessment.

The model has not previously taken any account of farm escapees in other areas. The incidence of farmed fish in catches at West Greenland and in most homewater fisheries has been estimated to be less than $1.5 \%$ (Table 3.3.7.1)(Hansen et al., 1997). These fish will therefore have a minimal effect on the PFA and National Conservation Limit assessments.

However, substantial numbers of farm escapees occur in coastal, fjord and river fisheries in Norway. These fish have previously been included in the run-reconstruction model and therefore contributed to the back-calculated estimate of recruitment (PFA), despite the fact that they are clearly not recruits from wild spawners but are derived from cage
production. The farm escapees have also been incorporated into the estimates of the numbers of spawners and thus contribute to the estimated egg-deposition. However, farm fish probably don't spawn as successfully as wild fish and their offspring may not be as viable as wild offspring; their contribution to the egg deposition should therefore be reduced.

The PFA and National Conservation Limits for Norway estimated using the unmodified NEAC PFA-CL model have therefore been compared with the same outputs from a modified version of the model taking account of the escapees. In the revised model, the numbers of wild fish of sea age ' i ' killed (including unreported catches) $\left(\mathrm{H}_{\mathrm{i}}\right)$, is estimated by adding a new parameter $\left(\mathrm{F}_{\mathrm{c}}\right)$ in the model. Thus:

$$
\mathrm{H}_{\mathrm{i}}=\left(\mathrm{C}_{\mathrm{i}} /\left(1-\mathrm{R}_{\mathrm{i}}\right) / \mathrm{U}_{\mathrm{i}}\right) *\left(1-\mathrm{F}_{\mathrm{c}}\right)
$$

where: $\mathrm{C}=$ catch of salmon in numbers;
$\mathrm{R}=$ estimated proportion of the total catch that is unreported;
$\mathrm{U}=$ average level of exploitation of the salmon stock;
$\mathrm{F}_{\mathrm{c}}=$ proportion of farmed fish in catches.

The PFA of wild salmon is estimated from the number of returns by back-calculating from the returning stock $\left(\mathrm{H}_{\mathrm{i}}\right)$.
In estimating the National Conservation Limit for Norway, the number of wild spawners was estimated by deducting the catch of wild fish $\left(C *\left(1-\mathrm{F}_{\mathrm{c}}\right)\right)$ from the estimated number of wild returns $\left(\mathrm{H}_{\mathrm{i}}\right)$. The total numbers of spawners was then calculated by raising this value for the proportion of farmed fish in samples taken on the spawning grounds ( $\mathrm{F}_{\mathrm{s}}$ ). However, since farmed fish are believed to be less successful spawners than wild fish their contribution must be weighted by a factor, W. Thus the effective number of spawners, $\mathrm{S}_{\mathrm{e}}$, is estimated as follows:

$$
\mathrm{S}_{\mathrm{e}}=\mathrm{H}_{\mathrm{i}}-\left(\mathrm{C} *\left(1-\mathrm{F}_{\mathrm{c}}\right) *\left(1 /\left(1-\mathrm{F}_{\mathrm{s}} * \mathrm{~W}\right)\right)\right)
$$

The effects of these changes on the assessment are shown in Table 3.6.6.1. Over the past 10 years the average proportions of farmed fish in catches of 1SW and MSW salmon ( $\mathrm{F}_{\mathrm{c}}$ ) have been $8 \%$ and $14 \%$ respectively. Removing these fish from the PFA assessment results in the estimated recruitment of 1SW and MSW salmon being reduced by an average of $10 \%$ and $18 \%$ over this period. The average proportion ( $\mathrm{F}_{\mathrm{s}}$ ) of farmed fish on the spawning grounds has been $11 \%$ and $15 \%$ for 1 SW and MSW salmon respectively. Taking account of these fish in the assessment, and assuming an arbitrary weighting factor (W) of 0.5 , the effective spawning numbers (and hence egg deposition) for 1SW and MSW salmon are reduced by $2 \%$ and $4 \%$ respectively. The conservation limit estimated by the national lagged-egg deposition model in this scenario is also reduced by about $4 \%$.

This model takes no account of the possibility of farmed fish having an adverse effect on the spawning success of wild fish. There is also considerable uncertainty about the spawning success of farmed escapees (factor 'W'); the assessment currently undertaken by the Working Group assumes that farmed and wild fish contribute equally to the egg deposition. If farmed fish make less contribution to the egg deposition (W closer to 1 ), the effective spawning escapement (and egg deposition) will be reduced further.

For the 2001 assessment, the Norwegian input data for the PFA model have been split into three regions. However, it has not been possible to provide data on farm escapees for these regions. As a result the potential errors discussed above must be taken into account when considering the catch advice.

### 3.6.7 Sensitivity analysis of the PFA model

A sensitivity analysis for the spreadsheet model which generates pre-fishery abundance (PFA) estimates based upon a run-reconstruction analysis for national salmon stocks in the North East Atlantic Commission Area (NEAC) of NASCO is described by Potter, et al. (1998).

The model (as applied to individual national data sets) is relatively insensitive to the ranges of natural mortality, M, and the time between recruitment and return to home waters, $t$, estimates currently used. Although ' $t$ ' can be estimated quite accurately, there is considerable uncertainty about the true values of M at different stages in the marine phase of the life
cycle. As suggested in Section 2.6, if estimates of M are revised in the future this parameter may be sensitive to changes and affect the assessments more than is indicated in the present analysis. (This issue is being investigated further as a separate task with the EU SALMODEL Concerted Action programme.) Variation in the values of non-reporting rates, ' $R$ ', and exploitation rates, ' $U$ ', have a greater effect on the estimates of PFA. The PFA estimate is most sensitive to ' $R$ ' when it is large and to ' $U$ ', when this is small. The latter is a particular concern because exploitation rates are being reduced in many countries.

The sensitivity of the overall assessment of PFA for the NEAC Area and for the Northern and Southern European stock complexes will therefore depend on the values of the various parameters provided for different countries, but these will also be weighted by the national catches. It is thus not immediately apparent to which parameter values the assessment will be most sensitive. A spreadsheet has been therefore been prepared which evaluates the effects on the overall assessment of changing individual parameter values (Tables 3.6.7.1). The evaluation is based upon the data inputs used for the PFA assessment this year (2001).

Table 3.6.7.1 shows the effects ( $\%$ change) on the assessment of PFA of maturing and non-maturing 1SW salmon from the Northern and Southern Europe of making the following changes to individual national or regional parameter values:

- adding 0.1 (10\%) to ' R '
- $\quad$ adding 0.1 ( $10 \%$ ) to ' $U$ '
- adding 2 months to ' t '
- multiplying ' $R$ ' by 1.2
- multiplying 'U' by 1.2
- multiplying ' $t$ ' by 1.2
[Adding 0.1 to parameters tends to weight the effects on low values, whereas multiplying them by 1.2 weights that effects on larger values.]

At this level of disaggregation the model is fairly sensitive to some parameter values. Changes (as described above) to the eight parameter values listed in the text table below have a greater than $5 \%$ effect on the respective PFA estimates (Table 3.6.7.1); several have effects of more than $10 \%$.

| Rank | Country | Sea-age | Parameter |
| :---: | :--- | :--- | :--- |
| 1 | Russia | MSW | non-reporting rate |
| 2 | Scotland (E) | MSW | exploitation rate |
| 3 | Russia | MSW | exploitation rate |
| 4 | Scotland (W) | MSW | exploitation rate |
| 5 | Scotland (E) | 1SW | exploitation rate |
| 6 | Ireland | 1SW | non-reporting rate |
| 7 | Ireland | 1SW | exploitation rate |
| 8 | Scotland (E) | MSW | non-reporting rate |

This analysis does not provide any indication of the reliability of any of the parameter values currently used in the NEAC PFA assessment. It merely indicates the effects on the PFA estimates when modest changes (adding 0.1 or multiplying by 1.2 ) are made to individual values. However, the sensitivity analysis indicates that particular attention should be paid to ensuring that the parameter values listed in the text table below are accurate.

### 3.7 Development of age-specific conservation limits

### 3.7.1 Progress with setting river-specific conservation limits

These issues are being dealt with by the EU funded Concerted Action entitled
"A co-ordinated approach towards the development of a scientific basis for management of wild Atlantic salmon in the North-East Atlantic - SALMODEL" (Contract No: QLK5-CT1999-01546). A summary taken from the first years report of this CA is given below:

While NASCO's remit in distant water fisheries requires an international approach, the use of conservation limits at national, regional and local levels is also highly important. At these levels, data on compliance with conservation limits for individual rivers or groups of rivers provides important data on status of stocks. These data are in some cases already being used to manage fisheries at regional and local levels and this is expected to increase as more river-specific conservation limits are set. The use of river specific conservation limits is now generally (though not universally) accepted as providing the most viable means of providing management advice for salmon at all levels from river through to stock complexes. Delivery of conservation limits at all these levels can be enhanced through international cooperation and sharing of data and techniques.

In all, there are around 25 stock and recruitment datasets in the NEAC area, ranging from long time series to rivers where $\mathrm{S} / \mathrm{R}$ relationships are in the process of being (or could be) developed. These include a mixture of smaller rivers and tributaries of large river systems. Given the time and resource difficulties with collecting meaningful $S / R$ data, it is unlikely that further datasets will be developed in the near future. However, as these rivers are spread throughout the NEAC area and cover a wide array of river types (apart from larger river systems) and productivity levels, even incomplete $\mathrm{S} / \mathrm{R}$ datasets may provide useful information for helping transport conservation limits to rivers with little or no data (i.e. conservation limits could be tuned to different productivity levels, by comparison across $\mathrm{S} / \mathrm{R}$ datasets, even if the local data are only partial).

As noted above, two countries (UK(England \& Wales ) and France) are already using river specific conservation limits to derive national conservation limits. These are supplemented by UK(N. Ireland) and Ireland, where river-specific conservation limits have been produced, but, as these are still viewed as preliminary they have not yet been used for inclusion in the ICES catch advice process. A majority of countries surveyed in the NEAC area (5) have still to develop even interim conservation limits for their rivers, although most are actively working towards this. For example, Sweden has indicated a target for production of river-specific conservation limits by 2001.

The rate of development of river-specific conservation limits has reflected inter alia difficulties across various countries with availability and representativeness of $S / R$ data, together with the logistical difficulty of accurately surveying large numbers of rivers in remote locations. As a result, only 271 out of a total of 1446 rivers (18.7\%) have river-specific conservation limits at present, and many of those are at interim/developmental stages.

One tendency is to use the most comprehensive $\mathrm{S} / \mathrm{R}$ data sets and transport conservation limits from these to many rivers. For example, the R. Bush $\mathrm{S} / \mathrm{R}$ dataset has been used by UK(N. Ireland), UK (England \& Wales) and Ireland to derive river-specific conservation limits for most of their rivers. This raises issues of differences in production across these rivers, together with a critical dependence on the value of a single $S / R$ dataset if there is dynamic change in that stock. In contrast, in Norway a good $\mathrm{S} / \mathrm{R}$ dataset exists for one small river in the south of the country (River Imsa), but this has not been used to set river-specific conservation limits for the whole country, because of concerns about representativeness of the data.

An intermediate approach is the French example, where a river in Brittany with good $\mathrm{S} / \mathrm{R}$ data has been used primarily for setting Biological Reference Points (BRP) in that region, but also in other regions after correction/calibration for local productivity, based on restricted local information in some cases. Similarly, in England \& Wales, the Environment Agency adjusts $S / R$ relationships according to a simple model of river type (productivity) and make further adjustments to take account of local knowledge of river structure, productivity and stock characteristics, where known.

The challenge is to take what is really a continuum of possible approaches to identifying and transporting BRPs and providing guidelines that can suggest a best approach for particular situations, taking into account limitations of the data and the methodology.

The process of setting river-specific conservation limits depends on being able to transport BRPs to other rivers . There are a variety of transport methods in use in the NEAC area, all based on measuring some attributes of area of productive habitat and relative levels of productivity in other rivers. These range from remote sensing (e.g. aerial photographs), through map-based measurements (e.g. catchment area/ gradient/wetted area) to in-river surveys of productive habitat area. In practical terms, remote sensing alone is unlikely to provide satisfactory resolution for meaningful transport, while logistical/resource difficulties mean that in-river surveys of all rivers will be impossible. A trend is emerging for top down surveys (incorporated into GIS supported production models), with in-river surveys used to provide ground truthing and calibration. SALMODEL will address development of survey and transport techniques so that where possible common standards can be applied in future.

While summing river-specific conservation limits is preferable to the pseudo stock-recruitment method for setting national conservation limits, efforts have been made to improve the data inputs used in the latter approach, which will continue to be used until sufficient progress has been made the individual river assessments.

Legitimate concerns have been expressed about the suitability of using single conservation limits for management of larger river systems known to have genetically differentiated sub-populations at the sub-catchment level. A related concern highlighted the possible impact of a single river conservation limit on sub-populations having different production characteristics.

Despite the problems, several countries have developed river-specific conservation limits for all their salmon rivers, though the methods used for transport vary considerably. In one or two cases, river specific conservation limits have been set based on the crudest level of transport (catchment area) and therefore these are not yet regarded as reliable enough to replace the ICES pseudo $\mathrm{S} / \mathrm{R}$ data for setting national conservation limits (Section 3.7.2). While it is noted that NASCO has specifically asked for the development of age specific conservation limits, there has been little progress with dividing river specific conservation limits between sea-age groups.

### 3.7.2 Changes to the National Conservation Limits model

As indicated above, relatively few river specific conservation limits have been developed for salmon stocks in the NEAC area. An interim approach has therefore been developed for estimating national conservation limits for countries that cannot provide an estimate based upon river specific estimates. The approach is based on establishing quasi-stockrecruitment relationships for national salmon stocks in the North East Atlantic Commission (NEAC) area (Potter et al., 1998).

In brief, the model provides a means for relating the estimates of numbers of spawners and recruits derived from the PFA model. This is addressed by converting the numbers of 1SW and MSW spawners into numbers of eggs deposited, using the proportion of female fish in each age class and the average number of eggs produced per female. The egg deposition in year ' $n$ ' is assumed to contribute to the recruitment in years ' $n+3$ ' to ' $n+8$ ' in proportion to the numbers of smolts produced of ages 1 to 6 years, and these proportions are therefore used to estimate the 'lagged egg deposition' contributing to the recruitment of maturing and non-maturing 1SW fish in year ' $n+8$ '. The plots of lagged eggs (stock) against the 1SW adults in the sea (recruits) have been presented as 'pseudo-stock-recruitment' relationships.

ICES and NASCO currently define the conservation limit for salmon as the stock size that will result in the maximum sustainable yield in the long term (i.e. $\mathrm{S}_{\mathrm{MSY}}$ ). However, it is not straightforward to estimate this point on the national stock-recruitment relationships because the replacement line is not known. (The replacement line is the line on which 'stock' equals 'recruits'.) This is the case for the quasi-stock-recruitment relationships established by the national model because the stock is expressed as eggs while the recruits are expressed as adult salmon. The Working Group has previously used the following three non-parametric methods (ICES (1993/Assess:10)) to provide Options for setting the conservation limits:

Option 1: the minimum stock size previously observed

Option 2: the stock size where the 90th \%ile of survival intersects the 90th \%ile of recruitment
Option 3: the stock size where the 90th \%ile of survival intersects the median recruitment level.

These values have then been evaluated on the basis that the conservation limit should be set at the egg-deposition level below which recruitment begins to decline, and that if no decline in recruitment is apparent over the range of eggdeposition previously observed, the conservation limit should be set at the lowest previously recorded egg-deposition.

The Working Group considered a more objective approach for setting a conservation limit from the quasi-stockrecruitment relationships. The approach assumes that there is a critical point below which recruitment decreases with stock towards zero levels of stock and recruitment, and above which recruitment is constant.
i.e. $\quad R=\beta \times S \quad S<$ Change point

$$
R=\mu_{S} \quad S \geq \text { Change point }
$$

where

$$
\beta=\frac{\mu_{S}}{\text { Change point }} \text {. }
$$

This is a non-linear model with parameters $\mu_{\mathrm{S}}$ and Change point, estimated by minimising the residual sum of squares. The minimum is obtained by searching for the Change point over the range $[0, \max (S)]$. The procedure is demonstrated for two simulated data sets:

$$
\begin{array}{ll}
R=S_{l}+\text { noise } & S_{l}=5,6, \ldots 14 \\
R=S_{2}+\text { noise } & S_{2}=15,16, \ldots 25
\end{array}
$$

The results are shown in Figure 3.7.2.1. The top pair of figures correspond to a low level of variability, and show on the left, the residual sum of squares for different values of the Change point, passing through a well defined minimum at Change point $=15.8$. The figure on the right shows the data with the fitted model corresponding to the minimum residual sum of squares. The bottom pair of figures correspond to a higher level of superimposed variability, and we see that the optimum value of Change point (14.0) is less well defined. It should be noted that in both cases that the residual sum of squares is not a smooth function, and there are discontinuities as the Change point coincides with each data point.

This approach was applied to the 2000 national stock-recruitment relationships and in most cases the model provided a conservation limit very close to the Option (1 to 3) previously selected by the Working Group. The Working Group therefore concluded that, being more objective, this approach was more appropriate for the evaluation of the national conservation limits.

### 3.7.3 National Conservation Limits

The national model has been run for the countries for which no river specific conservation limits have been developed (i.e. all countries except France, UK(England \& Wales) and Sweden). The outputs are illustrated in Appendix 8. For Iceland, UK(Northern Ireland) and UK(Scotland) the input data for the PFA analysis (1971-2000) have been provided for two regions; the lagged spawner analysis has therefore been conducted for each region separately and the estimated conservation limits summed for the country. The conservation limits derived from the national model and river specific estimates are shown in Table 3.7.3.1. The Working Group has previously noted that outputs from the national model are only designed to provide a rough guide to the status of stocks in the NEAC area, where more reliable data are not available because the approach depends upon a number of assumptions. One feature is that it tends to provide estimates close to the minimum level of recruitment previously seen (as estimated by the PFA model). In some cases this will be a conservative estimate of the true conservation limit ( $\mathrm{S}_{\mathrm{MSY}}$ ). It will also be noted that the conservation limit estimates may alter from year to year as the input of new data affects the 'pseudo-stock-recruitment relationship'. This further emphasises the fact that this approach only provides a basis for qualitative catch advice.

The estimated national conservation limits have been summed for Northern and Southern Europe (Table 3.7.3.1) and are given on Figures 3.6.4.4 and 3.6.4.6 for comparison with the estimated spawning escapement. The conservation limits have also been used to estimate the spawner escapement reserves (SERs) (i.e. the CL increased to take account of natural mortality between the recruitment date ( $1^{\text {st }} \mathrm{Jan}$ ) and return to home waters) for maturing and non-maturing 1SW salmon from the Northern and Southern Europe. The SERs are shown as horizontal lines in Figures 3.6.4.5 and 3.6.4.3. Evaluation of stocks against SERs is thought to be inappropriate for the total NEAC data (Figure 3.6.4.1). The Working Group also consider the current SER levels may be less appropriate for evaluating the historic status of stocks (e.g. pre1985), which in many cases have been estimated with less precision.

### 3.8 Catch options or alternative management advice

The Working Group has been asked to provide catch options or alternative management advice with an assessment of risks relative to the objective of exceeding stock conservation limits in the NEAC area. The Working Group reiterated its concerns about harvesting salmon in mixed stock fisheries, particularly for fisheries exploiting individual river stocks and sub-river populations that are at unsatisfactorily low levels. Annual adjustments in quotas or effort regulations based on changes in the mean status of the stocks is unlikely to provide adequate protection to the individual river stocks that are most heavily exploited by the fishery or are in the weakest condition.

The Working Group also emphasized that the national stock conservation limits discussed above are not appropriate for the management of homewater fisheries, particularly where these exploit separate river stocks. This is both because of
the relative imprecision of the national conservation limits and because they will not take account of differences in the status of different river stocks or sub-river populations. Nevertheless, the Working Group agreed that the combined conservation limits for the main stock groups (national stocks) exploited by the distant water fisheries could be used to provide general management advice for these fisheries.
In view of the uncertainties expressed about the most appropriate stock groupings and the preliminary nature of the conservation limit estimates, the Working Group is unable to provide quantitative catch options at this stage. In the absence of a predictive estimate of PFA and more reliable estimates of conservation limits, it is unlikely that quantitative catch advice will be developed in the near future. However, the Working Group feels that the following qualitative catch advice is appropriate based upon the PFA data and estimated SERs shown in Figures 3.6.4.3 and 3.6.4.5.

The Southern European stock complex is believed to include the main European stocks that have contributed fish to the West Greenland fishery; evidence from tagging studies suggests that the Nordic countries contribute relatively few fish to this fishery. It is therefore appropriate that the European input to the advice on the West Greenland fishery should be based principally on the status of non-maturing 1SW from the Southern area.

Provision of catch advice for the Faroes fishery is more complex. Recent tagging studies at Faroes (1991/92 1994/95), suggest that the main country contributing to the MSW salmon to the fishery is Norway, with significant contributions also from Scotland and Russia (Hansen and Jabobsen, 2001). The 1SW salmon caught in the fishery come mainly from the Southern European countries (Jacobsen et al. 2001). This therefore means that the catch advice for both Northern and Southern European stocks must be taken into account when considering management actions for the Faroes fishery.

For all fisheries, the Working Group considers that management of single stock fisheries should be based upon local assessments of the status of stocks. Conservation would be best achieved by fisheries in estuaries and rivers targeting stocks which have been shown to be above biologically-based escapement requirements.
[NB In the evaluation of the status of stocks, PFA or recruitment values should be assessed against the spawner escapement reserve values while the spawner numbers should be compared with the conservation limits.]

Northern European 1SW stocks: The spawning escapement of 1SW salmon from the Northern European stock complex have been within but close to safe biological limits in recent years, although there is evidence of an upturn in the past few years. It should be noted that the inclusion of farmed fish in the Norwegian data will result in the exploitable surplus being overestimated. The Working Group considers that overall exploitation of the stock complex at the current rate is acceptable, although the status of individual stocks varies considerably. Since very few of these salmon have been caught outside homewater fisheries in Europe, even when fisheries were operating in the Norwegian Sea, management of maturing 1SW salmon should be based upon local assessments of the status of river or sub-river stocks.

Northern European MSW stocks: The PFA of non-maturing 1SW salmon from Northern Europe has been declining since the mid 1980s and the exploitable surplus has fallen from around 1 million recruits in the 1970s to about half this level in recent years. The Working Group considers the Northern European MSW stock complex to be within safe biological limits, although it is recognised that the status of individual stocks will vary considerably. In addition, the inclusion of farmed fish in the Norwegian data will result in the exploitable surplus being overestimated. The Working Group therefore considers that great caution should be exercised in the management of these stocks particularly in mixed stock fisheries and exploitation should not be permitted to increase.

Southern European 1SW stocks: The spawning escapement for the whole stock complex has fallen below the conservation limit throughout the past 10 years. Moreover, recruitment of maturing 1SW salmon in the Southern European stock complex has been below any previously observed value throughout this period. In both 1999 and 2000 recruitment before exploitation was below the spawning escapement reserve. The Working Group considers that reductions in exploitation rates are required for as many stocks as possible and that mixed stock fisheries present particular threats to conservation.

Southern European MSW stocks: The PFA of non-maturing 1SW salmon from Southern Europe has been declining steadily since the 1970s and the spawning escapement for the whole stock complex has been close to or outside safe biological limits throughout much of this period. The upper $95 \%$ confidence limit for PFA of spawners has been below the spawner escapement reserve for the past four years. Qualitative projection of these estimates suggests that the PFA is likely to remain below this reserve in 2001. The Working Group considers that further reductions in exploitation rates are urgently required for as many stocks as possible and that mixed stock fisheries present particular threats to conservation.

Since 1995, regular post-smolt surveys have been conducted in the Norwegian Sea and adjacent areas during the summer months. These have been carried out by Norwegian research vessels during routine surface trawl surveys for pelagic species. Extra flotation on trawl-wings and on the head-rope were used when trawling for smolts. In the early 1990s, occasional post-smolt captures were also recorded when the pelagic trawl was being developed and tested, Since 1990 more than 2,400 post-smolts and 99 salmon have been caught in 1,863 surface trawls.

In 2000, trawling for post-smolts was carried out on seven cruises, four of which were specifically aimed at salmon using a 'Fish Lifter' for live fish sampling attached to the cod end of the trawl, as described in Holm et al. (2000) and Holst and Mc Donald (2000). One short cruise to the Voering Plateau was carried out during mid- June where, based on previous observations of the spatial and temporal distribution of post-smolts, it was hypothesized that an aggregation of post-smolts should occur. A special cruise was also conducted in July to four of the large northernmost fjords (Alta, Porsanger, Tana and Neiden) into which some of Norway's most important salmon rivers discharge. A further limited survey was also carried out to the southern edge of the North Atlantic Current where this flows eastwards into the Barents Sea. The southern Barents Sea is believed to be an important feeding area in the early marine phase for northern Norwegian and Russian salmon stocks. The aim of this cruise was to investigate the migration paths of the northern post-smolts, and the level of salmon lice infestations in these northern stocks; some of which are located in fjords (Tana and Neiden) without any fish farming activity.

Overall, more than 760 post-smolts and 38 salmon were captured during the 2000 surveys, the majority ( $\sim 60 \%$ ) in fjords or in coastal areas (Table 3.9.1). The total post-smolt distribution in 2000 is presented in Figure 3.9.1. Figures 3.9.2 and 3.9.3. provide detailed distributions of the catch locations and the numbers captured on the southwest coast of Norway and in the southern Barents Sea, respectively. The captures in both areas are distributed along the Norwegian Coastal Current. For the first time Carlin tagged fish from Ims, southwest Norway, and microtagged post-smolts, presumably of southern European origin, were taken in the same catch. Nine microtags and three Carlin-tags were recorded in a catch of 170 post-smolts and an additional three microtags were recorded in a catch of 34 post-smolts from a neighbouring area (Figures 3.9.1 and 2.4.4.1) This incidence supports the view that some of the post-smolts captured in these areas in 1990-1999 (2-year smolts, aged by scale reading) may have been of south- or southwestNorwegian origin. It is speculated that these fish would migrate along this route with the Atlantic current. The possibility of a common feeding area, explaining the observed correlation in growth and survival between the North Esk fish in Scotland and the SW- Norwegian Figgjo stock (Friedland et al. 1998), may also be supported by this finding. Scale readings from a sample of smolts from the largest catch in June, correspond to earlier recorded age distributions from this area, as the majority of the fish entered the sea as 2-year-old smolts or younger.

The high number of microtags taken in only a few hauls, suggest that even a modest, but co-ordinated effort to microtag more smolts in NE-Atlantic countries could add much to the knowledge about migration routes, spatial and temporal distribution and other characteristics of various salmon stocks in the sea.

The post-smolt CPUE values have been high in the Norwegian Sea during the last two years, assumed to reflect better timing of the cruises. It has previously been suggested that the trawling speed was too low for capturing larger salmon. This was also observed on video recordings from within the trawl during 2000 (Holm, pers obs.). Larger salmon were seen holding station close to the cod-end for a while, but were only occasionally recorded in the catches. Consequently, no efforts have been made to calculate CPUE for larger salmon. The high CPUE values for post-smolts give rise to concern about the interception of these fish in pelagic fisheries in international waters. This issue is addressed in Section 2.4.4.

### 3.10 Data deficiences and research needs in the NEAC Area

1. More research into the biology of salmon in the marine phase is required. This includes the need to monitor trends in marine mortality for a wider range of stocks than at present, and identify causes for mortality. It should also include the examination of relationships between postsmolt growth and marine mortality. The use of data storage tags will significantly improve the information on the marine life history of salmon.
2. Research on post-smolts in the early marine phase should be continued and expanded. This should include studies of interactions with parasites and assessments of the impact of sea lice on post-smolts.
3. A Study Group is required to quantitatively assess the level of bycatches of post-smolts in pelagic fisheries. It is recommended that such a group should comprise both those with information relating to postsmolt distribution and those who can provide information on the activity and distribution of pelagic fisheries.
4. A coordinated programme of tagging and release of farmed salmon should be undertaken to improve knowledge on the marine survival and migratory behaviour of these fish

If the commercial fishery at Faroes recommences, it is recommended that biological samples from the salmon caught should be collected. Historical samples from this fishery which have not yet been worked up should continue to be analysed.

Table 3.3.3.1 Numbers of gear units licensed or authorised by country and gear type.

| Year | England \& |  | Wales |  |  | UK (Scotland) |  | UK (N. Ireland) |  |  | Norway |  |  | $\begin{gathered} \hline \text { Driftnet } \\ \text { (No. nets) } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Gillnet licences | Sweepnet | $\begin{gathered} \text { Hand-held } \\ \text { net } \\ \hline \end{gathered}$ | Fixed engine | $\begin{gathered} \text { Rod \& } \\ \text { Line }^{1} \\ \hline \end{gathered}$ | Fixed engine ${ }^{2}$ | $\begin{gathered} \text { Net and } \\ \text { coble }{ }^{3} \\ \hline \end{gathered}$ | Driftnet | Draftnet | $\begin{gathered} \text { Bagnets } \\ \text { and boxes } \\ \hline \end{gathered}$ | Bagnet | Bendnet | Liftnet |  |
| 1966 | - | - | - | - |  | 3,513 | 861 | - | - | - | 7,101 | - | 55 | - |
| 1967 | - | - | - | - | - | 2,982 | 836 | - | - | - | 7,106 | 2,827 | 48 | 11,498 |
| 1968 | - | - | - | - | - | 3,495 | 970 | - | - | - | 6,588 | 2,613 | 36 | 9,149 |
| 1969 | - | - | - | - | - | 3,239 | 849 | 139 | 311 | 17 | 6,012 | 2,756 | 32 | 8,956 |
| 1970 | - | - | - | - | - | 2,861 | 775 | 138 | 306 | 17 | 5,476 | 2,548 | 32 | 7,932 |
| 1971 | - | - | - | - | - | 3,069 | 802 | 142 | 305 | 18 | 4,608 | 2,421 | 26 | 8,976 |
| 1972 | - | - | - | - | - | 3,437 | 810 | 130 | 307 | 18 | 4,215 | 2,367 | 24 | 13,448 |
| 1973 | - | - | - | - | - | 3,241 | 884 | 130 | 303 | 20 | 4,047 | 2,996 | 32 | 18,616 |
| 1974 | - | - | - | - | - | 3,182 | 777 | 129 | 307 | 18 | 3,382 | 3,342 | 29 | 14,078 |
| 1975 | - | - | - | - | - | 2,978 | 768 | 127 | 314 | 20 | 3,150 | 3,549 | 25 | 15,968 |
| 1976 | - | - | - | - | - | 2,854 | 756 | 126 | 287 | 18 | 2,569 | 3,890 | 22 | 17,794 |
| 1977 | - | - | - | - | - | 2,742 | 677 | 126 | 293 | 19 | 2,680 | 4,047 | 26 | 30,201 |
| 1978 | - | - | - | - | - | 2,572 | 691 | 126 | 284 | 18 | 1,980 | 3,976 | 12 | 23,301 |
| 1979 | - | - | - | - | - | 2,698 | 747 | 126 | 274 | 20 | 1,835 | 5,001 | 17 | 23,989 |
| 1980 | - | - | - | - | - | 2,892 | 670 | 125 | 258 | 20 | 2,118 | 4,922 | 20 | 25,652 |
| 1981 | - | - | - | - | - | 2,704 | 647 | 123 | 239 | 19 | 2,060 | 5,546 | 19 | 24,081 |
| 1982 | - | - | - | - | - | 2,377 | 641 | 123 | 221 | 18 | 1,843 | 5,217 | 27 | 22,520 |
| 1983 | 232 | 209 | 333 | 74 | - | 2,514 | 659 | 120 | 207 | 17 | 1,735 | 5,428 | 21 | 21,813 |
| 1984 | 226 | 223 | 354 | 74 | - | 2,438 | 630 | 121 | 192 | 19 | 1,697 | 5,386 | 35 | 21,210 |
| 1985 | 223 | 230 | 375 | 69 | - | 1,999 | 524 | 122 | 168 | 19 | 1,726 | 5,848 | 34 | 20,329 |
| 1986 | 220 | 221 | 368 | 64 | - | 1,976 | 583 | 121 | 148 | 18 | 1,630 | 5,979 | 14 | 17,945 |
| 1987 | 213 | 206 | 352 | 68 | - | 1,693 | 571 | 120 | 119 | 18 | 1,422 | 6,060 | 13 | 17,234 |
| 1988 | 210 | 212 | 284 | 70 | - | 1,536 | 390 | 115 | 113 | 18 | 1,322 | 5,702 | 11 | 15,532 |
| 1989 | 201 | 199 | 282 | 75 | - | 1,224 | 347 | 117 | 108 | 19 | 1,888 | 4,100 | 16 | 0 |
| 1990 | 200 | 204 | 292 | 69 | - | 1,276 | 334 | 114 | 106 | 17 | 2,375 | 3,890 | 7 | 0 |
| 1991 | 199 | 187 | 264 | 66 | - | 1,144 | 306 | 118 | 102 | 18 | 2,343 | 3,628 | 8 | 0 |
| 1992 | 203 | 158 | 267 | 65 | - | 857 | 296 | 121 | 91 | 19 | 2,268 | 3,342 | 5 | 0 |
| 1993 | 187 | 151 | 259 | 55 | - | 909 | 266 | 120 | 73 | 18 | 2,869 | 2,783 | - | 0 |
| 1994 | 177 | 158 | 257 | 53 | 293,759 | 753 | 245 | 119 | 68 | 18 | 2,630 | 2,825 | - | 0 |
| 1995 | 163 | 156 | 249 | 47 | 243,288 | 737 | 226 | 122 | 68 | 16 | 2,542 | 2,715 | - | 0 |
| 1996 | 151 | 132 | 232 | 42 | 231,744 | 614 | 203 | 117 | 66 | 12 | 2,280 | 2,860 | - | 0 |
| 1997 | 139 | 131 | 231 | 35 | 269,705 | 671 | 196 | 116 | 63 | 12 | 2,002 | 1,075 | - | 0 |
| 1998 | 130 | 129 | 196 | 35 | 233,401 | 537 | 151 | 117 | 70 | 12 | 1,865 | 1,027 | - | 0 |
| 1999 | 120 | 109 | 178 | 30 | 185,502 | 355 | 109 | 113 | 52 | 11 | 1,649 | 989 | - | 0 |
| 2000 | 110 | 101 | 150 | 28 | 174,690 | 185 | 89 | 109 | 57 | 10 | 1.577 | 982 | - | 0 |
| Mean 1995-99 | 141 | 131 | 217 | 38 | 232728 | 583 | 177 | 117 | 64 | 13 | 2068 | 1733 |  |  |
| $\%$ change ${ }^{4}$ | -21.8 | -23.1 | -30.9 | -25.9 | -24.9 | -68.3 | -49.7 | -6.8 | -10.7 | -20.6 | -23.7 | -43.3 |  |  |
| Mean 1990-99 | 167 | 152 | 243 | 50 | 242900 | 785 | 233 | 118 | 76 | 15 | 2282 | 2513 |  |  |
| $\%$ change ${ }^{4}$ | -34.1 | -33.3 | -38.1 | -43.7 | -28.1 | -76.4 | -61.8 | -7.4 | -24.9 | -34.6 | -30.9 | -60.9 |  |  |

[^0]Table 3.3.3.1 continued Number of gear units licensed or authorised by country and gear type.

| Year | Ireland |  |  |  | Finland |  |  |  | France |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Driftnets No. | Draftnets | Other nets Commercial | Rod | The Teno River |  |  | R. Näätämö <br> Recreational <br> fishery | Rod and line licences | Com. nets in freshwater ${ }^{4}$ | Licences in estuary ${ }^{4,5}$ |
|  |  |  |  |  | Recreational fishery Tourist anglers |  | Local rod and net fishery Fishermen |  |  |  |  |
|  |  |  |  |  | Fishing days | Fishermen |  | Fishermen |  |  |  |
| 1966 | 510 | 742 | 214 | 11,621 | - | - | - | - | - | - | - |
| 1967 | 531 | 732 | 223 | 10,457 | - | - | - | - | - | - | - |
| 1968 | 505 | 681 | 219 | 9,615 | - | - | - | - | - | - | - |
| 1969 | 669 | 665 | 220 | 10,450 | - | - | - | - | - | - | - |
| 1970 | 817 | 667 | 241 | 11,181 | - | - | - | - | - | - | - |
| 1971 | 916 | 697 | 213 | 10,566 | - | - | - | - | - | - | - |
| 1972 | 1,156 | 678 | 197 | 9,612 | - | - | - | - | - | - | - |
| 1973 | 1,112 | 713 | 224 | 11,660 | - | - | - | - | - | - | - |
| 1974 | 1,048 | 681 | 211 | 12,845 | - | - | - | - | - | - | - |
| 1975 | 1,046 | 672 | 212 | 13,142 | - | - | - | - | - | - | - |
| 1976 | 1,047 | 677 | 225 | 14,139 | - | - | - | - | - | - | - |
| 1977 | 997 | 650 | 211 | 11,721 | - | - | - | - | - | - | - |
| 1978 | 1,007 | 608 | 209 | 13,327 | - | - | - | - | - | - | - |
| 1979 | 924 | 657 | 240 | 12,726 | - | - | - | - | - | - | - |
| 1980 | 959 | 601 | 195 | 15,864 | - | - | - | - | - | - | - |
| 1981 | 878 | 601 | 195 | 15,519 | 16,859 | 5,742 | 677 | 467 | - | - | - |
| 1982 | 830 | 560 | 192 | 15,697 | 19,690 | 7,002 | 693 | 484 | 4,145 | 55 | 82 |
| 1983 | 801 | 526 | 190 | 16,737 | 20,363 | 7,053 | 740 | 587 | 3,856 | 49 | 82 |
| 1984 | 819 | 515 | 194 | 14,878 | 21,149 | 7,665 | 737 | 677 | 3,911 | 42 | 82 |
| 1985 | 827 | 526 | 190 | 15,929 | 21,742 | 7,575 | 740 | 866 | 4,443 | 40 | 82 |
| 1986 | 768 | 507 | 183 | 17,977 | 21,482 | 7,404 | 702 | 691 | 5,919 | $58{ }^{1}$ | 86 |
| 1987 | - | - | - | - | 22,487 | 7,759 | 754 | 689 | 5,804 ${ }^{1}$ | $87^{2}$ | 80 |
| 1988 | 836 | - | - | 11,539 | 21,708 | 7,755 | 741 | 538 | 4,413 | 101 | 76 |
| 1989 | 801 | - | - | 16,484 | 24,118 | 8,681 | 742 | 696 | 3,826 | 83 | 78 |
| 1990 | 756 | 525 | 189 | 15,395 | 19,596 | 7,677 | 728 | 614 | 2,977 | 71 | 76 |
| 1991 | 707 | 504 | 182 | 15,178 | 22,922 | 8,286 | 734 | 718 | 2,760 | 78 | 71 |
| 1992 | 691 | 535 | 183 | 20,263 | 26,748 | 9,058 | 749 | 875 | 2,160 | 57 | 71 |
| 1993 | 673 | 457 | 161 | 23,875 | 29,461 | 10,198 | 755 | 705 | 2,111 | 53 | 55 |
| 1994 | 732 | 494 | 176 | 24,988 | 26,517 | 8,985 | 751 | 671 | 1,680 | 17 | 59 |
| 1995 | 768 | 512 | 164 | 27,056 | 24,951 | 8,141 | 687 | 716 | 1,881 | 17 | 59 |
| 1996 | 778 | 523 | 170 | 29,759 | 17,625 | 5,743 | 672 | 814 | 1,806 | 21 | 69 |
| 1997 | 852 | 531 | 172 | 31,873 | 16,255 | 5,036 | 616 | 588 | 2,974 | 10 | 59 |
| 1998 | 874 | 513 | 174 | 31,565 | 18,700 | 5,759 | 621 | 673 | 2,358 | 16 | 63 |
| 1999 | 874 | 499 | 162 | 32,493 | 22,935 | 6,857 | 616 | 850 | 2,232 | 15 | 61 |
| 2000 | 871 | 490 | 158 | 33,527 | 28.385 | 8.275 | 633 | 624 | $2.745^{3}$ | 16 | 35 |
| Mean 1995-99 | 829 | 516 | 168 | 30549 | 20093 | 6307 | 642 | 728 | 2250 | 16 | 62 |
| $\%$ change $^{6}$ | 5.0 | -5.0 | -6.2 | 9.7 | 41.3 | 31.2 | -1.5 | -14.3 | 22.0 | 0.0 | -43.7 |
| Mean 1990-99 | 771 | 509 | 173 | 25245 | 22571 | 7574 | 693 | 722 | 2294 | 36 | 64 |
| $\% \text { change }^{6}$ | 13.0 | -3.8 | -8.8 | 32.8 | 25.8 | 9.3 | -8.6 | -13.6 | 19.7 | -54.9 | -45.6 |

${ }^{1}$ Common licence for salmon and seatrout introduced in 1986 leading to a short-term increase in the number of licences issued.
Since 1987 fishermen have been obliged to declere their catch
This figure is an estimate from a sample of anglers, the sea trout and salmon angling licenses being common since 2000
${ }^{4}$ The number of licences, 1999 included, indicates only the number of fishermen (or boats allowed to fish for salmon. It overestimates the actual number of fishermen fishing for salmon up to 2 or 3 times.
Adour estuary only southwest of France.
${ }^{6}(2000 /$ mean - $) * 100$

Table 3.3.4.1 Nominal catch of SALMON in NEAC Area (in tonnes round fresh weight), 1960-2000

| Year | Homewater countries | Faroes <br> (1) | Other catches in international waters | Total <br> Reported Catch | Unreported catches |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | $\begin{gathered} \text { NEAC } \\ \text { Area } \\ \hline \end{gathered}$ | International waters (2) |
| 1960 | 5540 | - | - | 5540 | - | - |
| 1961 | 4753 | - | - | 4753 | - | - |
| 1962 | 6709 | - | - | 6709 | - | - |
| 1963 | 6276 | - | - | 6276 | - | - |
| 1964 | 7150 | - | - | 7150 | - | - |
| 1965 | 6456 | - | - | 6456 | - | - |
| 1966 | 6052 | - | - | 6052 | - | - |
| 1967 | 7526 | - | - | 7526 | - | - |
| 1968 | 6146 | 5 | 403 | 6554 | - | - |
| 1969 | 6281 | 7 | 893 | 7181 | - | - |
| 1970 | 5882 | 12 | 922 | 6816 | - | - |
| 1971 | 5582 | - | 471 | 6053 | - | - |
| 1972 | 6597 | 9 | 486 | 7092 | - | - |
| 1973 | 7331 | 28 | 533 | 7892 | - | - |
| 1974 | 7027 | 20 | 373 | 7420 | - | - |
| 1975 | 7116 | 28 | 475 | 7619 | - | - |
| 1976 | 5314 | 40 | 289 | 5643 | - | - |
| 1977 | 5209 | 40 | 192 | 5441 | - | - |
| 1978 | 4966 | 37 | 138 | 5141 | - | - |
| 1979 | 5121 | 119 | 193 | 5433 | - | - |
| 1980 | 5434 | 536 | 277 | 6247 | - | - |
| 1981 | 4909 | 1025 | 313 | 6247 | - | - |
| 1982 | 4471 | 606 | 437 | 5514 | - | - |
| 1983 | 5873 | 678 | 466 | 7017 | - | - |
| 1984 | 4769 | 628 | 101 | 5498 | - | - |
| 1985 | 5533 | 566 | - | 6099 | - | - |
| 1986 | 6183 | 530 | - | 6713 | - | - |
| 1987 | 4830 | 576 | - | 5406 | 2554 | - |
| 1988 | 5284 | 243 | - | 5527 | 3087 | - |
| 1989 | 4059 | 364 | - | 4423 | 2103 | - |
| 1990 | 3420 | 315 | - | 3735 | 1779 | 180-350 |
| 1991 | 2822 | 95 | - | 2917 | 1555 | 25-100 |
| 1992 | 3343 | 23 | - | 3366 | 1825 | 25-100 |
| 1993 | 3311 | 23 | - | 3334 | 1471 | 25-100 |
| 1994 | 3563 | 6 | - | 3569 | 1157 | 25-100 |
| 1995 | 3277 | 5 | - | 3282 | 942 | n/a |
| 1996 | 2750 | - | - | 2750 | 947 | $\mathrm{n} / \mathrm{a}$ |
| 1997 | 2074 | - | - | 2074 | 827 | n/a |
| 1998 | 2222 | 6 | - | 2228 | 1108 | n/a |
| 1999 | 2074 | - | - | 2074 | 877 | $\mathrm{n} / \mathrm{a}$ |
| 2000 | 2633 | - | - | 2633 | 1135 | $\mathrm{n} / \mathrm{a}$ |
| Means |  |  |  |  |  |  |
| 1995-1999 | 2479 | 6 | - | 2482 | 940 | - |

Table 3.3.5.1 CPUE for salmon rod fisheries in Finland (Teno, Naatamo), France, the River Bush (UK(N.Ireland)).

|  | Finland (Teno River) |  | Finland (Naatamo River) |  | France | UK(N.Ire.)(R.Bush) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Catch per angler season | Catch per angler day | Catch per angler season | Catch per angler day | Catch per angler season | Catch per rod day |
| Year | kg | kg | kg | kg | Number | Number |
| 1974 |  | 2.8 |  |  |  |  |
| 1975 |  | 2.7 |  |  |  |  |
| 1976 |  | - |  |  |  |  |
| 1977 |  | 1.4 |  |  |  |  |
| 1978 |  | 1.1 |  |  |  |  |
| 1979 |  | 0.9 |  |  |  |  |
| 1980 |  | 1.1 |  |  |  |  |
| 1981 | 3.2 | 1.2 |  |  |  |  |
| 1982 | 3.4 | 1.1 |  |  |  |  |
| 1983 | 3.4 | 1.2 |  |  |  | 0.248 |
| 1984 | 2.2 | 0.8 | 0.5 | 0.2 |  | 0.083 |
| 1985 | 2.7 | 0.9 | n/a | n /a |  | 0.283 |
| 1986 | 2.1 | 0.7 | $\mathrm{n} / \mathrm{a}$ | n/a |  | 0.274 |
| 1987 | 2.3 | 0.8 | n/a | n/a | 0.39 | 0.194 |
| 1988 | 1.9 | 0.7 | 0.5 | 0.2 | 0.73 | 0.165 |
| 1989 | 2.2 | 0.8 | 1.0 | 0.4 | 0.55 | 0.135 |
| 1990 | 2.8 | 1.1 | 0.7 | 0.3 | 0.71 | 0.247 |
| 1991 | 3.4 | 1.2 | 1.3 | 0.5 | 0.60 | 0.396 |
| 1992 | 4.5 | 1.5 | 1.4 | 0.3 | 0.94 | 0.258 |
| 1993 | 3.9 | 1.3 | 0.4 | 0.2 | 0.88 | 0.341 |
| 1994 | 2.4 | 0.8 | 0.6 | 0.2 | 2.31 | 0.205 |
| 1995 | 2.7 | 0.9 | 0.5 | 0.1 | 1.15 | 0.206 |
| 1996 | 3.0 | 1.0 | 0.7 | 0.2 | 1.57 | 0.267 |
| 1997 | 3.4 | 1.0 | 1.1 | 0.2 | $0.43{ }^{1}$ | 0.338 |
| 1998 | 3.0 | 0.9 | 1.3 | 0.3 | 0.67 | 0.569 |
| 1999 | 3.7 | 1.1 | 0.8 | 0.2 | 0.76 | 0.273 |
| 2000 | 5.0 | 1.5 | 0.9 | 0.2 | 0.79 | 0.259 |
| Mean |  |  |  |  |  |  |
| 1995-99 | 3.2 | 1.0 | 0.9 | 0.2 | 0.90 | 0.331 |

${ }^{1}$ Large numbers of new, inexperienced anglers in 1997 because cheaper licence types were introduced.

Table 3.3.5.2 CPUE for salmon rod fisheries in the Barent Sea and White Sea basin in Russia.

| Barents Sea Basin, catch per angler day |  |  |  |  | White Sea Basin, catch per angler day |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Rynda | Kharlovka | Varzina | Iokanga | Ponoy | Varzuga | Kitsa | Umba |
| 1991 |  |  |  |  | 2.794 | 1.870 |  | 1.330 |
| 1992 | 2.370 | 1.454 | 1.070 | 0.135 | 3.489 | 2.261 | 1.209 | 1.366 |
| 1993 | 1.177 | 1.464 | 0.488 | 0.650 | 2.881 | 1.278 | 1.425 | 2.720 |
| 1994 | 0.710 | 0.847 | 0.548 | 0.325 | 2.332 | 1.596 | 1.588 | 1.436 |
| 1995 | 0.486 | 0.782 | 1.220 | 0.718 | 3.459 | 2.524 | 1.784 | 1.196 |
| 1996 | 0.703 | 0.845 | 1.502 | 1.398 | 3.503 | 1.444 | 1.761 | 0.930 |
| 1997 | 1.197 | 0.709 | 0.613 | 1.411 | 5.330 | 2.364 | 2.482 | 1.457 |
| 1998 | 1.010 | 0.551 | 0.441 | 0.868 | 4.544 | 2.284 | 2.784 | 0.979 |
| 1999 | 0.947 | 0.642 | 0.427 | 1.193 | 3.300 | 1.710 | 1.657 | 0.756 |
| 2000 | 1.348 | 0.769 | 0.565 | 2.283 | 3.494 | 1.526 | 3.018 | 1.245 |
| Mean |  |  |  |  |  |  |  |  |
| 1995-99 | 0.869 | 0.706 | 0.841 | 1.118 | 4.027 | 2.065 | 2.094 | 1.064 |

Table 3.3.5.3 CPUE data for net and fixed engine salmon fisheries by Region in UK (England and Wales), 1988-2000. (Data expressed as catch per licence-day and catch per licence-tide for Midlands, Wales and North West.)

| Year | Region |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | North East | Southern | South West | Midlands ${ }^{1,2}$ | Wales ${ }^{2}$ | North West ${ }^{2}$ |
| 1988 | 5.49 | 10.15 |  |  | - | - |
| 1989 | 4.39 | 16.80 |  |  | 0.90 | 0.82 |
| 1990 | 5.53 | 8.56 |  |  | 0.78 | 0.63 |
| 1991 | 3.20 | 6.40 |  |  | 0.62 | 0.51 |
| 1992 | 3.83 | 5.00 |  |  | 0.69 | 0.40 |
| 1993 | 6.43 | No fishing |  |  | 0.68 | 0.63 |
| 1994 | 7.53 | - |  |  | 1.02 | 0.71 |
| 1995 | 7.84 | - |  |  | 1.00 | 0.79 |
| 1996 | 3.74 | - |  |  | 0.73 | 0.59 |
| 1997 | 5.30 | - | 0.59 |  | 0.77 | 0.35 |
| 1998 | 5.12 | - | 0.78 | 0.25 | 0.69 | 0.32 |
| 1999 | 7.28 | - | 0.67 | 0.36 | 0.83 | 0.37 |
| 2000 | 11.10 | - | 0.96 | 0.43 | 0.40 | 0.64 |
| Mean |  |  |  |  |  |  |
| 1995-99 | 5.86 | - |  |  | 0.80 | 0.48 |

[^1]Table 3.3.5.4 CPUE data for Scottish net fisheries.
Catch in numbers of fish per unit effort.

| Year | Fixed engine | Net and coble CPUE |
| :---: | :---: | :---: |
|  | Catch/trap month ${ }^{1}$ | Catch/crew month |
| 1952 | 33.91 | 156.39 |
| 1953 | 33.12 | 121.73 |
| 1954 | 29.33 | 162.00 |
| 1955 | 37.09 | 201.76 |
| 1956 | 25.71 | 117.48 |
| 1957 | 32.58 | 178.70 |
| 1958 | 48.36 | 170.39 |
| 1959 | 33.30 | 159.34 |
| 1960 | 30.67 | 177.80 |
| 1961 | 31.00 | 155.17 |
| 1962 | 43.89 | 242.00 |
| 1963 | 44.25 | 182.86 |
| 1964 | 57.92 | 247.11 |
| 1965 | 43.67 | 188.61 |
| 1966 | 44.86 | 210.59 |
| 1967 | 72.57 | 329.80 |
| 1968 | 46.99 | 198.47 |
| 1969 | 65.51 | 327.64 |
| 1970 | 50.28 | 241.91 |
| 1971 | 57.19 | 231.61 |
| 1972 | 57.49 | 248.04 |
| 1973 | 73.74 | 240.60 |
| 1974 | 63.42 | 257.11 |
| 1975 | 53.63 | 235.71 |
| 1976 | 42.88 | 150.79 |
| 1977 | 45.58 | 188.67 |
| 1978 | 53.93 | 196.07 |
| 1979 | 42.20 | 157.19 |
| 1980 | 37.65 | 158.62 |
| 1981 | 49.60 | 183.86 |
| 1982 | 62.26 | 181.89 |
| 1983 | 56.20 | 206.83 |
| 1984 | 58.98 | 160.98 |
| 1985 | 54.48 | 156.55 |
| 1986 | 75.93 | 204.87 |
| 1987 | 64.34 | 147.14 |
| 1988 | 51.91 | 204.53 |
| 1989 | 71.68 | 268.78 |
| 1990 | 33.31 | 148.37 |
| 1991 | 35.62 | 100.44 |
| 1992 | 59.10 | 151.85 |
| 1993 | 52.29 | 124.06 |
| 1994 | 93.23 | 123.40 |
| 1995 | 75.03 | 139.72 |
| 1996 | 60.51 | 110.93 |
| 1997 | 33.95 | 56.27 |
| 1998 | 36.75 | 65.54 |
| 1999 | 24.30 | 69.70 |
| 2000 | 38.80 | 86.90 |
| Mean |  |  |
| 1995-99 | 46.11 | 88.43 |

Table 3.3.5.5 Catch per unit effort for the marine fishery in Norway. The CPUE is expressed as number of salmon caught per net day in Bagnets and Bendnets devided by salmon wight.

|  | Bagnet |  |  |  | Bendnet |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | $<\mathbf{3 k g}$ | $\mathbf{3 - 7} \mathbf{~ k g}$ | $\mathbf{> 7 ~ k g}$ |  | $<\mathbf{3 k g}$ | $\mathbf{3 - 7} \mathbf{~ k g}$ | $\mathbf{> 7 ~ k g}$ |
| 1998 | 0.88 | 0.66 | 0.12 |  | 0.80 | 0.56 | 0.13 |
| 1999 | 1.16 | 0.72 | 0.16 |  | 0.75 | 0.67 | 0.17 |
| 2000 | 2.01 | 0.90 | 0.17 |  | 1.24 | 0.87 | 0.17 |


| Table 3.3.5.6 |  |  | Fisheries in the North East Atlantic, summary of trend analyses based on non-parametric method (1000 iterations) (p $<0.1$ means significance upward trend, $\mathrm{p}>0.9$ means significant downward trend). |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Section/Data type | Life stage | Period (years) | Fisheries | ' p ' value | Trend |
| Section 3.3.5 |  |  |  |  |  |
| CPUE |  | 10 | UK (Scotland) net fisheries. Catch/trap month | 0.99 | Dn |
|  |  | 10 | UK (England \& Wales) net and fixed engines. Catch per licence-day | 0.07 | Up |
|  |  | 10 | Finland (Teno, Näätämö) and France. Rod catch/season, | 0.37 | Nt |
|  |  | 10 | Finland (Teno, Näätämö) and UK (N Ireland) (Bush). Rod catch/day | 0.72 | Nt |
|  |  | 9 | Russia (Barents Sea basin: Rynda, Kharlovka, Varzina, Iokanga). Rod catch/day | 0.83 | Nt |
|  |  | 10 | Russia (White Sea basin: Ponoy, Varzuga, Kitsa, Umba). Rod catch/day | 0.03 | Up |
| Section 3.3.9 |  |  |  |  |  |
| Exploitation rates | 1 SW | 10 | $\begin{aligned} & \text { Burrishoole + Delphi + Screebe + Shannon + Bunowen + Corrib } \\ & \text { (Ire), Dee (UK (E\&W)), North Esk (UK (Scot)), Bush (UK (NI)), } \\ & \text { Imsa + Drammen (Nor), Ellidaar + Blandar (Ice), Lagan (Swe) } \end{aligned}$ | 0.16 | Nt |
|  | 1 SW | 5 | ```Burrishoole + Delphi + Screebe + Shannon + Bunowen + Corrib (Ire), Dee (UK (E&W)), North Esk (UK (Scot)), Bush (UK (NI)), Imsa + Drammen (Nor), Ellidaar + Blandar + Vesturdalsa (Ice), Lagan (Swe)``` | 0.33 | Nt |
|  | 2 SW | 10 | Corrib (Irl), North Esk (UK Scot), Dee (UK (NI)), Imsa + Drammen (Nor), Blanda (Ice), Lagan (Swe) | 1 | Dn |
|  | 2 SW | 5 | Corrib (Irl), North Esk (UK Scot), Dee (UK (NI)), Imsa + Drammen (Nor), Blanda + Vesturdalsa (Ice), Lagan (Swe) | 0.86 | Nt |
|  | All ages | 10 | B.Z.Litsa, Ura, Tuloma, Kola (Russia, Barents Sea basin) | 0.99 | Dn |
|  |  | 5 |  | 0.46 | Nt |
|  | All ages | 10 | Ponoy, Kitsa,Varzuga, Umba (Russia, White Sea basin) | 0.99 | Dn |
|  |  | 5 |  | 0.94 | Dn |

Table 3.3.6.1 The percent of 1SW salmon in catches from countries in the North East Atlantic Commission, 1987-2000.

| Year | Finland | France | Iceland | Norway | Russia | Sweden | UK (Scot) | UK (E\&W) <br> (1) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 66 | 77 |  | 61 | 71 |  | 61 |  |
| 1988 | 63 | 29 |  | 64 | 53 |  | 57 |  |
| 1989 | 66 | 33 |  | 73 | 73 | 41 | 63 |  |
| 1990 | 64 | 45 |  | 68 | 73 | 70 | 48 |  |
| 1991 | 59 | 39 |  | 65 | 70 | 71 | 53 |  |
| 1992 | 70 | 48 |  | 62 | 72 | 68 | 55 | 77 |
| 1993 | 58 | 74 |  | 61 | 61 | 62 | 57 | 78 |
| 1994 | 55 | 55 | 64 | - | 69 | 65 | 54 | 77 |
| 1995 | 59 | 60 | 72 | 58 | 70 | 78 | 53 | 72 |
| 1996 | 80 | 51 | 74 | 53 | 80 | 63 | 54 | 65 |
| 1997 | 70 | 51 | 73 | 64 | 82 | 54 | 54 | 73 |
| 1998 | 75 | 71 | 82 | 66 | 82 | 59 | 58 | 83 |
| 1999 | 81 | 27 | 71 | 65 | 78 | 71 | 43 | 70 |
| 2000 | 67 | 58 | 84 | 67 | 75 | 69 | 48 | 79 |
| Means |  |  |  |  |  |  |  |  |
| 1995-99 | 73 | 52 | 74 | 62 | 78 | 65 | 52 | 73 |
| 1990-99 | 67 | 52 | 73 | 62 | 74 | 66 | 53 | 74 |

1. Refers to rod and line catches only.

Table 3.3.7.1 Estimated catches (in tonnes round fresh weight) of wild, farmed and ranched salmon in national catches in the North East Atlantic (figures for 2000 include provisional values).

| Country | Catches of Salmon |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Year/Seas on | Wild | FW Farmed | SEA <br> Farmed | Total <br> Farmed | Ranched | Total |
| Norway | 1989 | 707 | 29 | 166 | 195 | 3 | 905 |
|  | 1990 | 709.8 | 29 | 185 | 214 | 6.2 | 930 |
|  | 1991 | 682.5 | 20 | 169 | 189 | 5.5 | 877 |
|  | 1992 | 653.7 | 27 | 176 | 203 | 10.3 | 867 |
|  | 1993 | 707 | 18 | 191 | 209 | 7 | 923 |
|  | 1994 | 781 | 18 | 187 | 205 | 10 | 996 |
|  | 1995 | 654 | 13 | 170 | 183 | 2 | 839 |
|  | 1996 | 557 | 19 | 203 | 222 | 8 | 787 |
|  | 1997 | 430 | 21 | 177 | 198 | 2 | 630 |
|  | 1998 | 530 | 29 | 180 | 209 | 1 | 740 |
|  | 1999 | 612 | 20 | 178 | 198 | 1 | 811 |
|  | 2000 | 940 | 34 | 201 | 235 | 1 | 1176 |
| Faroes | 1990/1991 | 117.2 |  |  | 84.8 | 0 | 202 |
|  | 1991/1992 | 20.4 |  |  | 10.6 | 0 | 31 |
|  | 1992/1993 | 16.1 |  |  | 5.9 | 0 | 22 |
|  | 1993/1994 | 5.8 |  |  | 1.2 | 0 | 7 |
|  | 1994/1995 | 4.8 |  |  | 1.2 | 0 | 6 |
|  | 1995/1996 | 0.8 |  |  | 0.2 | 0 | 1 |
|  | $1996 / 1997$ | 0 |  |  | 0 | 0 | $0$ |
|  | $1997 / 1998^{5}$ | - |  |  | - | - | 6 |
|  | $1998 / 1999$ | 0 |  |  | 0 | 0 | 0 |
|  | $1999 / 2000^{5}$ | - |  |  | - | - | $8$ |
|  | $200072001$ | - |  |  | - | - | - |
| Finland | $1991$ | 68 |  |  | $<1$ | 0 | 69 |
|  | 1992 | 77 |  |  | $<1$ | 0 | 78 |
|  | 1993 | 70 |  |  | $<1$ | 0 | 70 |
|  | 1994 | 49 |  |  | $<1$ | 0 | 49 |
|  | $1995$ | 48 |  |  | $<1$ | $0$ | $48$ |
|  | 1996 | 44 |  |  | $<1$ | 0 | 44 |
|  | 1997 | 45 |  |  | $<1$ | 0 | 45 |
|  | 1998 | 48 |  |  | $<1$ | 0 | 48 |
|  | $1999$ | 63 |  |  | $<1$ | 0 | 63 |
|  | $2000$ | 95 |  |  | $<1$ | 0 | 95 |
| France | 1991 | 13 |  |  | 0 | 0 | 13 |
|  | 1992 | 20 |  |  | 0 | 0 | 20 |
|  | 1993 | 16 |  |  | 0 | 0 | 16 |
|  | 1994 | 18 |  |  | 0 | 0 | 18 |
|  | $1995$ | 9 |  |  | 0 | 0 | 9 |
|  | 1996 | 14 |  |  | 0 | 0 | 14 |
|  | 1997 | 8 |  |  | 0 | 0 | 8 |
|  | $1998$ | 9 |  |  | 0 | 0 | 9 |
|  | $1999$ | $11$ |  |  | 0 | $0$ | $11$ |
|  | 2000 | 11 |  |  | 0 | 0 | 11 |
| Iceland ${ }^{1}$ | 1991 | 130 |  |  | 3 | 345 | 478 |
|  | 1992 | 175 |  |  | + | 460 | 635 |
|  | 1993 | 160 |  |  | - | 496 | 656 |
|  | $1994$ | 140 |  |  | - | 308 | 448 |
|  | $1995$ | 150 |  |  | - | 298 | 448 |
|  | 1996 | 122 |  |  | - | 239 | 361 |
|  | 1997 | 106 |  |  | - | 50 | 156 |
|  | 1998 | 130 |  |  | - | 34 | 164 |
|  | 1999 | 119 |  |  | - | 26 | 145 |
|  | 2000 | 82 |  |  | - | 2 | 84 |
| Ireland ${ }^{2}$ | $1991$ | 400 |  |  | 1.7 | 2.3 | 404 |
|  | 1992 | 620 |  |  | 3.8 | 6.7 | 630 |
|  | 1993 | 531 |  |  | 1.9 | 8.1 | 541 |
|  | $1994$ | 789 |  |  | 2.6 | 12.5 | 804 |
|  | $1995$ | $774$ |  |  | $0.7$ | $14.8$ | 790 |
|  | 1996 | 668 |  |  | 0.7 | 15.9 | 685 |

Table 3.3.7.1 (continued). Estimated catches (in tonnes round fresh weight) of wild, farmed and ranched salmon in national catches in the North East Atlantic (figures for 2000 include provisional values).

| Country | Catches of Salmon |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Year/Season | Wild | FW Farmed | SEA <br> Farmed | Total Farmed | Ranched | Total |
| Ireland continued | 1997 | 565 |  |  | 1.7 | 2.9 | 570 |
|  | 1998 | 615 |  |  | 1.1 | 7.5 | 624 |
|  | 1999 | 509 |  |  | 2.1 | 3.6 | 515 |
|  | 2000 | 617 |  |  | 1.1 | 3.3 | 621 |
| Russia | 1991 | 215 |  |  | 0 | 0 | 215 |
|  | 1992 | 167 |  |  | 0 | 0 | 167 |
|  | 1993 | 139 |  |  | 0 | 0 | 139 |
|  | 1994 | 141 |  |  | 0 | 0 | 141 |
|  | 1995 | 128 |  |  | 0 | 0 | 128 |
|  | 1996 | 131 |  |  | 0 | 0 | 131 |
|  | 1997 | 111 |  |  | 0 | 0 | 111 |
|  | 1998 | 131 |  |  | 0 | 0 | 131 |
|  | 1999 | 102 |  |  | 0 | 0 | 102 |
|  | 2000 | 124 |  |  | 0 | 0 | 124 |
| Sweden | 1991 | 23 |  |  | 1 | $14^{3}$ | 38 |
|  | $1992$ | 24 |  |  | 1 | $24^{3}$ | 49 |
|  | 1993 | 35 |  |  | 1 | $20^{3}$ | 56 |
|  | 1994 | 15 |  |  | 1 | $29^{3}$ | 44 |
|  | 1995 | 12 |  |  | 1 | $24^{3}$ | 37 |
|  | $1996$ | 10 |  |  | 1 | $22^{3}$ | 33 |
|  | 1997 | 9 |  |  | 0 | $10^{3}$ | 19 |
|  | 1998 | 9 |  |  | 0 | $6^{3}$ | 15 |
|  | 1999 | 8 |  |  | 0 | $8^{3}$ | 16 |
|  | 2000 | 12 |  |  | 0 | $21^{3}$ | 33 |
| UK (E\&W) | 1991 | 200 |  |  | 0 | 0 | 200 |
|  | $1992$ | 186 |  |  | 0 | 0 | 186 |
|  | 1993 | 263 |  |  | 0 | 0 | 263 |
|  | 1994 | 307 |  |  | 0 | 0 | 307 |
|  | 1995 | 295 |  |  | 0 | 0 | 295 |
|  | 1996 | 183 |  |  | 0 | 0 | 180 |
|  | $1997$ | 142 |  |  | 0 | 0 | 142 |
|  | 1998 | 123 |  |  | 0 | 0 | 125 |
|  | 1999 | 150 |  |  | 0 | 0 | 152 |
|  | 2000 | 214 |  |  | 0 | 0 | 214 |
| UK (N.Ire) | 1991 | 54 |  |  | <1 | - | 55 |
|  | $1992$ | 85.3 |  |  | 1.1 | 2.6 | 89 |
|  | 1993 | 80.5 |  |  | 0.2 | 2.3 | 83 |
|  | 1994 | 90.1 |  |  | 0.5 | 0.4 | 91 |
|  | 1995 | 80.6 |  |  | 1.5 | 0.9 | 83 |
|  | 1996 | 74.7 |  |  | n/a | 2.3 | 77 |
|  | 1997 | 90.7 |  |  | 0.07 | 2.2 | 93 |
|  | 1998 | 76.6 |  |  | 0.03 | 1.0 | 78 |
|  | 1999 | 50.9 |  |  | 0.67 | 1.4 | 53 |
|  | 2000 | 75 |  |  | ?? | 2.5 | 78 |
| $\mathrm{UK}(\mathrm{Scot})^{4}$ | 1991 | 448 |  |  | 14 | 0 | 462 |
|  | $1992$ | 569 |  |  | 31 | 0 | 600 |
|  | 1993 | 516 |  |  | 31 | 0 | 547 |
|  | 1994 | 644 |  |  | 5 | 0 | 649 |
|  | 1995 | 586 |  |  | 2 | 0 | 588 |
|  | 1996 | 427 |  |  | $<1$ | 0 | 427 |
|  | 1997 | 296 |  |  | $<1$ | 0 | 296 |
|  | 1998 | 283 |  |  | <1 | 0 | 283 |
|  | 1999 | 198 |  |  | 1 | 0 | 199 |
|  | 2000 | 191 |  |  | 1 | 0 | 192 |

1. " + " indicates a small but unquantified catch. 2. Smolts released for enhancement of stocks or rod fisheries are categorised as wild
2. Fish released for mitigation purposes and not expected to contribute to natural spawning.
3. Data from 1994 onwards is the figure reported in national catch statistics, previous years' data have been calculated from a sampling programme.
4. Breakdown of the $1997 / 1998 \& 199 / 2000$ catches not
available.

Table 3.3.7.2 Proportion of farmed Atlantic salmon (unweighted means) in marine fisheries in Norway 1989-2000. $\mathrm{n}=$ number of salmon examined.

|  | Coast |  |  |  | Fjords |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | n | No. localities | \% | Range | n | No. localities | \% | Range |
| 1989 | 1217 | 7 | 45 | 7-66 | 803 | 4 | 14 | 8-29 |
| 1990 | 2481 | 9 | 48 | 16-64 | 940 | 5 | 15 | 6-36 |
| 1991* | 1245 | 6 | 49 | 29-63 | 336 | 3 | 10 | 6-16 |
| 1992 | 1162 | 7 | 44 | 4-72 | 307 | 1 | 21 | - |
| 1993 | 1477 | 7 | 47 | 1-60 | 520 | 4 | 20 | $7-47$ |
| 1994 | 1087 | 7 | 34 | 2-62 | 615 | 4 | 19 | 2-42 |
| 1995 | 976 | 7 | 42 | 2-57 | 745 | 4 | 17 | 2-47 |
| 1996* | 1183 | 6 | 54 | 35-68 | 678 | 4 | 16 | 3-22 |
| 1997 | 2046 | 8 | 47 | 7-68 | 793 | 5 | 42 | 15-85 |
| 1998 | 1194 | 8 | 45 | 6-61 | 1152 | 5 | 43 | 9-91 |
| 1999 | 1351 | 8 | 35 | 20-59 | 872 | 5 | 41 | $2-85$ |
| 2000 | 1996 | 9 | 31 | 8-68 | 1291 | 7 | 31 | 2-80 |

* In 1991 and 1996 the coastal results do not include the locality in Finnmark.

Table 3.3.7.3 Proportion of farmed Atlantic salmon (unweighted means) in rod catches (1 June-18 August) and brood stock catches (18 August-30 November) in Norway in 1989-2000. ( $\mathrm{n}=$ number of salmon examined; $\mathrm{R}=$ number of rivers sampled ).

| Year | 1 June-18 August |  |  |  | 18 August-30 November |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | n | R | \% | Range | n | R | \% | Range |
| 1989 | 5970 | 39 | 7 | 0-26 | 1892 | 19 | 35 | 2-77 |
| 1990 | 5380 | 39 | 7 | 0-55 | 2144 | 24 | 34 | 2-82 |
| 1991 | 4563 | 31 | 5 | 0-23 | 1799 | 26 | 24 | 0-82 |
| 1992 | 4259 | 32 | 5 | 0-24 | 1489 | 22 | 26 | 0-71 |
| 1993 | 4070 | 29 | 5 | 0-22 | 1213 | 21 | 22 | 0-75 |
| 1994 | 3243 | 18 | 4 | 0-19 | 1699 | 19 | 22 | 0-75 |
| 1995* | 3480 | 26 | 5 | 0-20 | 1279 | 19 | 29 | 0-71 |
| 1996* | 3020 | 29 | 7 | 0-54 | 1443 | 23 | 31 | 0-82 |
| 1997* | 2747 | 30 | 9 | 0-34 | 1892 | 36 | 29 | 0-83 |
| 1998* | 4161 | 33 | 9 | 0-46 | 1546 | 26 | 22 | 0-97 |
| 1999* | 5003 | 34 | 6 | 0-29 | 1755 | 23 | 15 | 0-53 |
| 2000* | 5036 | 32 | 7 | 0-48 | 1835 | 31 | 8 | 0-37 |

[^2]Table 3.3.7.4 Proportions of escaped farmed Atlantic salmon in 1987 - 2000 in the River Teno (Finland, Norway) during the fishing season (June-August) and after the season (September-October).

| Year | Fishing season (June-August) |  |  | After season (September-October) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | samples <br> (n) | farmed fish <br> (n) | farmed fish (\%) | samples <br> (n) | farmed fish <br> (n) | farmed fish (\%) |
| 1987 | 1430 | 1 | 0.07 |  |  |  |
| 1988 | 1026 | 1 | 0.10 |  |  |  |
| 1989 | 2096 | 5 | 0.24 |  |  |  |
| 1990 | 2467 | 11 | 0.45 | 19 | 10 | 47.3 |
| 1991 | 3146 | 11 | 0.35 | 7 | 4 | 37.5 |
| 1992 | 3748 | 2 | 0.05 |  |  |  |
| 1993 | 2413 | 1 | 0.04 |  |  |  |
| 1994 | 1529 | 6 | 0.39 |  |  |  |
| 1995 | 1604 | 5 | 0.31 |  |  |  |
| 1996 | 2173 | 3 | 0.14 | 8 | 1 | 12.5 |
| 1997 | 3881 | 7 | 0.18 | 28 | 0 | 0.0 |
| 1998 | 3722 | 10 | 0.27 |  |  |  |
| 1999 | 6243 | 10 | 0.16 |  |  |  |
| 2000 | 3448 | 5 | 0.15 |  |  |  |

Table 3.3.7.5 Salmon farm escapees in R. Bush (UK, N.Ireland) based on trapping of the total run throughout the year. (Note: 1994 data includes 14 escapees entering in January 1995).

|  |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{1 9 9 1}$ | $\mathbf{1 9 9 2}$ | $\mathbf{1 9 9 3}$ | $\mathbf{1 9 9 4}$ | $\mathbf{1 9 9 5}$ | $\mathbf{1 9 9 6}$ | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ |  |
| Total run |  |  |  |  |  |  |  |  |  |  |  |
| (excl. ranched) | 2344 | 2570 | 3253 | 2064 | 1527 | 1099 | 1681 | 2961 | 959 | 950 |  |
| No. escapees | 3 | 24 | 18 | 54 | 6 | 2 | 4 | 6 | 5 | 4 |  |
| \% in sample | 0.13 | 0.93 | 0.55 | 2.62 | 0.39 | 0.18 | 0.24 | 0.20 | 0.5 | 0.3 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |

Table 3.3.7.6 Geographical distribution by frequency (\%) of escaped farmed fish located among commercial catch samples for UK (Northern Ireland) and Ireland inshore catches (1991-2000).

|  |  | Frequency (\%) |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Location |  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| Northern <br> (UK) | Ireland | - | 3.72 | 0.26 | 1.18 | 4.03 | - | 0.14 | 0.2 | 1.9 | 3.1 |
| Donegal |  | 0.00 | 0.02 | 0.09 | 0.14 | 0.02 | 0.34 | 0.03 | 0.01 | 0.04 | 0.002 |
| Mayo | 1.16 | 1.69 | 0.27 | 0.10 | 0.14 | 0.25 | 0.27 | 0.17 | 0.79 | 0.62 |  |
| Galway | 0.39 | 0.10 | 0.06 | 0.08 | 0.03 | 0.00 | 0.06 | 0.10 | 0.51 | 2.16 |  |
| S. West | 0.00 | 0.01 | 1.05 | 1.08 | 0.19 | 0.42 | 0.47 | 1.10 | 0.69 | 0.05 |  |
| S. and East | - | - | - | - | - | 0.00 | - | - | - | 0.16 |  |

T able 3.3.7.7
Frequency of occurrence of escaped farmed salmon (detected by morphological characters and/or scales growth patterns) among Scottish fisheries for wild salmon (1981-2000).

| Year | Nets |  |  |  |  |  |  |  |  |  |  | Rods |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | East <br> Rigg <br> S <br> \% | Redpoint \% | Achilti buie \% | Culkein <br> Clachtol <br> \% | Strathy <br> \% | Bonar \% | Spey <br> \% | $\begin{gathered} \text { Dee } \\ \% \end{gathered}$ | N. Esk \% | $\begin{gathered} \text { Tay } \\ \% \end{gathered}$ | Tweed \% | Laxford \% | Annan \% | Cree \% | N. Esk \% | Dionard $\%$ |
| 1981 | 0 |  |  |  | 0 | 0 | 0 |  | 0 |  |  |  |  |  |  |  |
| 1982 | 0 |  |  |  | 0.3 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |
| 1983 | 0 |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |
| 1984 | 0 |  |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |
| 1985 | 0 |  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  | 0 |  |
| 1986 |  |  |  | 0.6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  | 0 |  |
| 1987 | 0 |  |  | 1.3 | 0 | 0 | 0 |  | 0 | 0 | 0 |  |  |  | 0 |  |
| 1988 |  |  |  | 1.5 | 0.6 | 0 | 0 |  | 0 | 0 | 0 |  |  |  | 0 |  |
| 1989 |  |  |  | 6.6 | 6.1 | 0.7 | 0.08 |  | 0 | 0 | 0 |  |  |  | 0 |  |
| 1990 |  | *22 |  | 4.7 | 3.8 | 0 | 0 |  | 0 | 0 | 0.13 |  |  |  | 0 |  |
| 1991 |  | 19.8 |  | 8.6 | 7.3 | 0.4 | 0.14 |  | 0.13 | 0 | 0 |  |  |  | 0 |  |
| 1992 |  | 18.5 |  | 3.5 | 2.3 | 0.5 | 0 |  | 0 | 0.13 | 0 |  |  |  | 0.16 |  |
| 1993 |  | 37.5 |  | 14.4 | 15.2 | 0.7 |  |  | 0 | 0 | 0 |  |  |  | 0.15 |  |
| 1994 |  |  |  | 7.7 | 7.1 | 0.6 |  |  | 0 | 0.18 | 0.4 |  |  |  | 0.3 |  |
| 1995 |  | 14.5 | 4.2 |  | 4.1 |  |  |  | 0 | 0 | 0 |  |  |  | 0 |  |
| 1996 |  | 4.84 | 6.9 |  | 3.4 |  |  |  | 0 | 0 | 0 |  |  |  | 0 |  |
| 1997 |  | 0 | 0 |  | 2.1 |  |  |  | 0 |  |  | 0.2 |  |  | 0 |  |
| 1998 |  |  | 3.45 | 2.8 | 0.5 |  |  |  | 0.05 |  | 0 | 0.0 |  |  | 0.35 |  |
| 1999 |  |  |  |  | 2.76 |  |  |  | 0.14 |  | 0 | 0.0 |  |  | 0 |  |
| 2000 |  |  |  |  | 2.6 |  |  |  | 0.05 |  | 0 | 9.4 | 0 | 1.9 | 0 | 0 |

* Carotenoid pigment analysis.

Table 3.3.9.1 Estimated exploitation rates (in \%) of salmon in homewater fisheries in the North East Atlantic area (Ireland)

|  | Ireland ${ }^{1}$ |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Lee | Shannon | CorCong | CorGal | Screebe | Delphi | Burrishoole | Bunowen | Eme | Moy |  |  |
|  | All types of gear |  |  |  |  |  |  |  |  |  | net | net |
| Year | HR | HR | HR | HR | HR | HR | HR | HR | HR | W | W | W |
|  | 1SW | 1SW | 1SW | 1SW | 1SW | 1SW | 1SW | 1SW | 1SW | 1SW | 1SW | 2SW |
| 1980 | 96 | 94 | 73 | - | - | - | 63 | - | - | - | 82 | 82 |
| 1981 | 91 | 96 | 53 | - | - | - | 85 | - | - | - | 95 | 45 |
| 1982 | 94 | 96 | 50 | 67 | - | - | 83 | - | - | - | 62 | 37 |
| 1983 | 84 | 97 | 67 | 84 | - | - | 84 | - | - | - | 59 | 31 |
| 1984 | 100 | 98 | 67 | 84 | 99 | - | 87 | - | 90 | - | 79 | 54 |
| 1985 | 98 | 98 | 44 | 50 | 100 | - | 86 | - | 96 | - | 68 | 37 |
| 1986 | 99 | 97 | - | 78 | 100 | - | 79 | - | 83 | - | - | - |
| 1987 | 92 | 98 | - | 59 | 100 | - | 74 | - | 85 | - | 54 | 56 |
| 1988 | 99 | 96 | 59 | - | 95 | - | 80 | - | 91 | - | 62 | 66 |
| 1989 | 88 | 94 | 53 | 76 | 96 | - | 76 | - | 85 | - | 54 | 23 |
| 1990 | 80 | 79 | 53 | 42 | 0 | - | 63 | - | 70 | 68 | 41 | 52 |
| 1991 | 99 | 86 | 60 | 68 | 96 | 92 | 83 | - | 94 | 85 | 40 | 60 |
| 1992 | - | 84 | 59 | 67 | 86 | 75 | 77 | 86 | - | 64 | 52 | - |
| 1993 | - | 89 | 62 |  | 98 | 74 | 73 | 90 | - | 95 | 60 | 16 |
| 1994 | - | 90 | - | 55 | 98 | 90 | 88 | 100 | - | - | 62 | 52 |
| 1995 | - | 91 | 57 | - | 93 | 76 | 58 | 99 | - | - | 66 | 100 |
| 1996 | - | 88 | - | - | 64 | 68 | 75 | 99 | - | - | 49 | 62 |
| 1997 | 90 | 83 | - | - | 78 | 68 | 68 | 94 | 83 | - | 60 | 16 |
| 1998 | 92 | 95 | 18 | 32 | 92 | 95 | 82 | 98 | 100 | - | 38 | 11 |
| 1999 | - | 100 | - | 20 | 67 | 79 | 85 | 100 | 79 | - | 43 | - |
| 2000 | $\mathrm{n} / \mathrm{a}$ | n/a | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | n/a | $\mathrm{n} / \mathrm{a}$ | n/a | n/a | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | n/a | $\mathrm{n} / \mathrm{a}$ |
| Mean |  |  |  |  |  |  |  |  |  |  |  |  |
| 1989-98 | 90 | 89 | 52 | 47 | 77 | 80 | 75 | 96 | 85 | 78 | 51 | 46 |
| 1994-98 | 91 | 91 | 38 | 26 | 79 | 77 | 74 | 98 | 87 |  | 51 | 47 |

[^3]Continued.....

Table 3.3.9.1 Estimated exploitation rates (in \%) of salmon in homewater fisheries in the North East
Atlantic area (UK)

| Year | UK (England and Wales) |  |  |  |  |  |  |  |  |  |  |  |  | UK (Northern Ireland) ${ }^{1}$ |  |  |  | UK (Scotland) ${ }^{2}$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sothern |  | Southwestern |  |  | Welsh |  |  |  | Northwestern |  |  |  |  |  |  |  |  |  |  |  |
|  | Test | Itchen | Frome | Tamar | Fowey | Dee | Dee | Dee | Taff | Leven | Kent | Lune | Lune | River Bush |  |  |  | North Esk |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  | net |  | net net |  |  |  |  |  |  |
|  | W/H | W | W | W | W | W | W | W | H | W | W | W | W | W | W/HR | HR1+ | HR2+ | W | W | W | W |
|  | (all ages) |  | (all ages) |  |  | 1SW | MSW | (all ages) |  | (all ages) |  |  |  | 1SW | 2SW | 1SW | 1SW | $1 \mathrm{SW}$ | $2 \mathrm{SW}$ | 1 SW | 2 SW |
| 1981 | - | - | - |  |  | - | - | - | - | - |  |  |  | - | - | - | - | 24 | 58 | 1 | 4 |
| 1982 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 30 | 48 | 1 | 9 |
| 1983 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 16 | 32 | 1 | 14 |
| 1984 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 29 | 42 | 2 | 15 |
| 1985 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 93 | - | 24 | 35 | 2 | 19 |
| 1986 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 82 | 75 | 43 | 30 | 3 | 21 |
| 1987 | - | - | - | - | - | - | - | - | - | - | - | - | - | 69 | 46 | 94 | 77 | 30 | 38 | 3 | 23 |
| 1988 | 39 | - | 10 | - | - | - | - | - | - | - | - | - | - | 65 | 36 | 72 | 57 | 35 | 37 | 3 | 27 |
| 1989 | 29 | 45 | 8 | - | - | - | - | - | - | - | - | - | 44 | 89 | 60 | 92 | 83 | 25 | 26 | 4 | 18 |
| 1990 | 36 | 51 | 11 | - | - | - | - | - | - | - | - | - | 36 | 61 | 38 | 63 | 70 | 36 | 34 | 6 | 29 |
| 1991 | 26 | 45 | 9 | - | - | - | - | - | - | - | - | - | 30 | 65 | 43 | 57 | 46 | 10 | 15 | 2 | 37 |
| 1992 | 25 | 27 | 11 | - | - | 14 | 18 | 15 | 5 | - | - | 23 | 30 | 56 | 33 | 74 | 75 | 28 | 27 | 7 | 28 |
| 1993 | 26 | 41 | 12 | - | - | 11 | 15 | 11 | 6 | - | - | 16 | 30 | 41 | 12 | 67 | 71 | 25 | 18 | 10 | 21 |
| 1994 | 25 | 44 | 14 | 13 | - | 15 | 21 | 22 | 5 | - | - | 29 | 35 |  | 40 | 71 | 64 | 18 | 18 | 5 | 24 |
| 1995 | 23 | 21 | 9 | 7 | 11 | 7 | 11 | 18 | 4 | 35 | 35 | 17 | 27 | 67 | 42 | 69 | - | 14 | 12 | 4 | 21 |
| 1996 | 19 | 47 | 13 | 7 | 8 | 9 | 11 | 17 | 3 | 21 | 31 | 17 | 24 |  | - | 81 | 77 | 19 | 10 | 4 | 32 |
| 1997 | 12 | 22 | 6 | 5 | 8 | 8 | 9 | 17 | 1 | 58 | 50 | 18 | 29 | 60 | - | 79 | 75 | 12 | 12 | 5 | 18 |
| 1998 | 18 | 18 | 6 | 6 | 7 | 10 | 10 | 15 | - | 43 | 65 | 12 | 14 | 26 | - | - | 32 | 23 | 12 | 5 | 20 |
| 1999 | 10 | 13 | n/a | 4 | 3 | 11 | 9 | 21 | 12 | 2 | 15 | 12 | 14 | 63 | - | 68 | 51 | 18 | 14 | 2 | 14 |
| $2000{ }^{3}$ | 9 | 9 | n/a | 4 | 5 | 6 | 17 | 14 | n/a | 2 | 16 | 8 | 15 | 55 | - | 75 | 67 | 33 | 24 | 1 | 20 |
| Mean |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1990-99 | 22 | 33 | 10 | 7 | 7 | 10 | 13 | 17 | 5 | - | - | 17 | 27 | 55 | 35 | 70 | 62 | 20 | 17 | 5 | 24 |
| 1995-99 | 16 | 24 | 09 | 06 | 07 | 09 | 10 | 18 | 05 | 32 | 39 | 15 | 22 | 54 | 42 | 74 | 59 | 17 | 12 | 04 | 21 |

${ }^{1}$ Estimate based on microtag recoveries raised to total catch and including estimate of non-catch fishing mortality.
${ }^{2}$ Estimate based on counter and catch figures.
${ }^{3}$ Provisional figures.
$\mathrm{HR}=$ Hatchery reared.
Table 3.3.9.1 continued........
$\mathrm{W}=$ Wild .
'-' = no data

Table 3.3.9.1 (cont'd) Estimated exploitation rates (in \%) of salmon in homewater fisheries in the North East Atlantic area (Iceland, Norway and Sweden)

| Year | Iceland ${ }^{1}$ |  |  |  |  | Norway ${ }^{2}$ |  |  |  |  |  |  |  |  |  | Sweden ${ }^{3}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ellidaar | Vesturdalsa |  | Blanda |  | Nausta | Øystenså | Orkla | Drammen |  |  | Imsa |  |  |  | Lagan |  |
|  | rod | rod rod |  | W | W | W | W | W | rod | net |  | net |  | net |  | net |  |
|  | W | W | W |  |  |  |  |  | W/HR | $\mathrm{HR}^{4}$ |  | W |  | $\mathrm{HR}^{4}$ |  | $\mathrm{HR}^{4}$ |  |
|  | 1SW | 1 SW | 2SW | 1 SW | 2 SW | All ages |  |  |  | 1SW | 2SW | 1SW | 2SW | 1SW | 2SW | 1SW | 2SW |
| 1985 | 40 | - | - | - | - | - | - | - | 33 | 57 | - | 73 | 94 | 81 | 100 | 81 | - |
| 1986 | 34 | - | - | 74 | 78 | - | - | - | 50 | 81 | 50 | 79 | 82 | 78 | 90 | 93 | 82 |
| 1987 | 54 | - | - | 75 | 72 | - | - | - | 44 | 64 | 52 | 56 | 95 | 83 | 95 | 78 | 55 |
| 1988 | 45 | - | - | 81 | 84 | - | - | - | 53 | 70 | 47 | 51 | 80 | 78 | 91 | 73 | 91 |
| 1989 | 41 | - | - | 72 | 73 | - | - | - | 35 | 40 | 59 | 65 | 74 | 44 | 65 | 76 | 86 |
| 1990 | 41 | - | - | 81 | 81 | - | - | - | 33 | 23 | 40 | 42 | 42 | 47 | 68 | 80 | 82 |
| 1991 | 37 | - | - | 75 | 86 | - | - | - | 28 | 54 | 59 | 37 | 72 | 50 | 66 | 91 | 92 |
| 1992 | 48 | - | - | 46 | 68 | - | - | - | 46 | - | 51 | 61 | 76 | 74 | 91 | 73 | 98 |
| 1993 | 41 | - | - | 57 | 64 | - | - | - | 45 | 20 | - | 53 | 80 | 85 | 89 | 89 | 82 |
| 1994 | 49 | - | - | 49 | 68 | - | - | 15 | 42 | 42 | 34 | 58 | 80 | 70 | 94 | 70 | 100 |
| 1995 | 43 | - | - | 53 | 90 | - | - | - | 53 | 29 | 40 | - | 86 | 56 | 88 | 58 | 70 |
| 1996 | 56 | 61 | 71 | 54 | 59 | - | - | 9 | 47 | 7 | 23 | 66 | - | 80 | 89 | 64 | 78 |
| 1997 | 50 | 60 | 80 | 40 | 63 | - | 9 | 12 | 44 | 15 | 23 | 58 | 80 | 67 | - | 55 | 58 |
| 1998 | 55 | 68 | 81 | 50 | 48 | 24 | 19 | 23 | 36 | 21 | 33 | 10 | 33 | 10 | 66 | 83 | 66 |
| 1999 | 48 | 55 | 69 | 40 | 55 | 16 | 22 | 22 | 42 | 5 | 0 | 0 | - | 19 | - | 49 | 17 |
| 2000 | 46 | 84 | 65 | 37 | 67 | 34 | 27 | 23 | 35 | 26 | 27 | 28 | - | 16 | 33 | 46 | - |
| Mean |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1990-99 | 47 | 61 | 75 | 55 | 68 | 20 | 17 | 16 | 42 | 24 | 34 | 43 | 69 | 56 | 81 | 71 | 74 |
| 1995-99 | 50 | 61 | 75 | 47 | 63 | 20 | 17 | 17 | 44 | 15 | 24 | 34 | 66 | 46 | 81 | 62 | 58 |

Estimate based on counter and catch figures
Estimates based on counter catch figures
${ }^{3}$ Estimate based on external tag recoveries and before 1994 on assumed $50 \%$
exploitation in the river brood stock fishery and in 1994-96 on mark-recovery estimates.
${ }^{4}$ HR in R. Drammen, R. Imsa and R. Lagan are pooled groups of $1+$ and $2+$ smolts.
${ }^{5}$ Provisional figures.
${ }^{6}$ Net only.
$\mathrm{W}=\mathrm{W}$ ild
$H R=$ Hatchery reared.
'-' = no data

Reporting rates for external tags:

|  |  |
| :--- | :--- |
| Norway | 0.50 |
| Sweden | 0.65 |

Sweden
Elsewhere
0.50

Continued...........

Table 3.3.9.1 (cont'd) Estimated exploitation rates (in \%) of salmon in homewater fisheries in the North East Atlantic (Sweden and Russia)


[^4]Table 3.4.1.1 Conservation limits achievement (egg deposition/conservation limit) in rivers in the NEAC area. (Rivers ranked by mean \% conservation Limit achieved over the last 10 years).

| Country | River | $\begin{aligned} & \text { O} \\ & 0 \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { } \\ & \text { O } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \underset{\sim}{\circ} \\ & \Gamma \end{aligned}$ | $\begin{aligned} & \infty \\ & \infty \\ & \text { o } \\ & \hline \end{aligned}$ | $\begin{aligned} & +\infty \\ & \text { on } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { ம } \\ & \text { © } \\ & \stackrel{-}{2} \\ & \hline \end{aligned}$ | $$ | $\begin{aligned} & \hat{\infty} \\ & \text { on } \\ & \hline \end{aligned}$ | $\begin{aligned} & \infty \\ & \infty \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { O} \\ & \mathbf{\infty} \\ & \hline- \\ & \hline \end{aligned}$ | $\begin{aligned} & 8 \\ & \stackrel{8}{7} \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { 항 } \\ & \text { 아 } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \underset{\sim}{8} \end{aligned}$ | $\begin{aligned} & \text { প } \\ & \text { or } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { অ } \\ & \stackrel{\rightharpoonup}{2} \\ & \hline \end{aligned}$ | $\stackrel{1}{8}$ | $\circ$ <br> $\stackrel{\circ}{8}$ <br> $\stackrel{\circ}{-}$ | $\begin{aligned} & \text { A} \\ & \text { or } \\ & \hline \end{aligned}$ | $\infty$ <br> $\stackrel{\circ}{\circ}$ | $\begin{aligned} & \hline 8 \\ & 8 \\ & \hline \end{aligned}$ | - |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| UK (E\&W) | Test | 4 | 5 |  |  |  |  |  |  |  | 3 | 4 | 4 | 4 | 3 | 4 | 4 | 4 | 5 | 3 | 3 | 4 |
| Russia | Litsa |  |  | 3 | 3 | 2 | 2 | 3 | 2 | 5 | 3 | 3 | 2 | 3 | 4 | 2 | 3 | 4 | 3 | 3 | 3 | 4 |
| UK (E\&W) | Lune |  |  |  |  |  |  |  |  |  | 3 | 3 | 3 | 4 | 3 | 3 | 3 | 3 | 3 | 2 | 3 | 2 |
| Ireland | Burrishoole | $\begin{aligned} & 2 \\ & 3 \end{aligned}$ | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 2 | 4 | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 3 | 3 |
| UK ( N I) | Bush |  | 3 | 3 | 3 | 3 | 2 | 1 | 2 | 2 | 4 | 2 | 2 | 2 | 2 | 3 | 3 | 3 | 3 | 2 | 4 | 4 |
| UK (E\&W) | Dee |  |  |  |  |  |  |  |  |  |  |  |  | 3 | 2 | 3 | 3 | 3 | 3 | 2 | 3 | 3 |
| Russia | Ura | 3 | 3 | 2 | 2 | 1 | 1 | 2 | 2 | 2 | 2 | 1 | 2 | 2 | 2 | 3 | 3 | 4 | 3 | 3 | 3 | 5 |
| Russia | Tuloma |  | 4 | 4 | 2 | 1 | 2 | 2 | 3 | 3 | 2 | 2 | 2 | 2 | 2 | 3 | 3 | 3 | 3 | 3 | 2 | 2 |
| UK (E\&W) | Coquet |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 1 | 2 | 2 | 2 | 2 | 2 |
| Russia | Varzuga |  | 2 | 2 | 3 | 2 | 2 | 2 | 2 | 1 | 1 | 1 | 2 | 1 | 2 | 3 | 2 | 3 | 3 | 2 | 2 | 2 |
| Russia | Kitsa | 2 | 3 | 2 | 2 | 3 | 2 | 1 | 2 | 1 | 2 | 2 | 2 | 3 | 1 | 1 | 3 | 2 | 2 | 2 | 1 | 1 |
| UK (E\&W) | Tamar |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 1 | 2 | 2 | 2 | 2 | 2 |
| UK (S) | North Esk |  | 1 | 1 | 1 | 2 | 1 | 1 | 2 | 1 | 1 | 2 | 1 | 1 | 1 | 1 | 1 | 2 | 1 | 1 | 1 | 1 |
| Farance | Nivelle |  |  |  |  | 3 | 4 | 2 | 2 | 2 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 2 | 2 | 1 |  |


| Percent of conservation attained: | $>200$ | $\mathbf{1}$ |
| :--- | ---: | :--- |
| (egg deposition/conservation limit) | $100-200$ | $\mathbf{2}$ |
|  | $50-100$ | $\mathbf{3}$ |
|  | $25-50$ | $\mathbf{4}$ |
|  | $0-25$ | $\mathbf{5}$ |
|  | N/A |  |

Table 3.4.2.1 Wild smolt counts and estimates, and juvenile survey data on various index streams in the North East atlantic (Finland, Norway and Sweden).


[^5]Continued.....

Table 3.4.2.2
Status of stocks in the North East Atlantic. Summary of trend analyses on smolt counts and survival based on a non-parametric method (1000 iterations). ( $\mathrm{p}<0.1$ means significance up) trend, $\mathrm{p}>0.9$ means significant downward trend).

| Type of data | Life stage | Period (years) | Rivers (Countries) | 'p' value | Trend |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Section 3.4.2 <br> Smolt counts | Smolts | 10 | Southern rivers: Oir (Fra), Burrishoole (Irl), Bush (UK NI), North Esk, Girnock, Baddoch (UK Scot) | 0.34 | Nt |
|  | Smolts | 5 |  | 0.14 | Nt |
|  | Smolts | 10 | Northern rivers: Orkla, Halselva, Stjordalselva (Nor), Högvadsån (Swe) | 0.99 | Dn |
|  | Smolts | 5 |  | 0.64 | Nt |
|  | Smolts | 10 | Southern + Northern rivers + Ellidaar, Vesturdalsa (Ice). | 0.98 | Dn |
|  | Smolts | 5 |  | 0.28 | Nt |
| Section 3.4.4 <br> Wild smolt survival | 1SW return to homewaters | 10 | Corrib (Irl), Bush (UK NI), Imsa (Nor), North Esk (UK Scot), Ellidaar + Midfjardara (Ice) | 0.99 | Dn |
|  | 1SW return to homewaters | 5 | Corrib (Irl)+Bush, Imsa (Nor), North Esk (UK Scot), Ellidaar+Midfjardara (Ice) | 0.95 | Dn |
|  | 2SW return to homewaters | 10 | Corrib (Irl), Imsa (Nor), North Esk (UK Scot), Midfjardara (Ice) | 0.48 | Nt |
|  | 2SW return to homewaters | 5 | Corrib (Irl), Imsa (Nor), North Esk (UK Scot), Midfjardara (Ice) | 0.11 | Nt |
| Section 3.4.4 <br> Hatchery smolt survival | 1SW return to homewaters | 10 | Shannon, Screebe, Delphi, Bunowen and Burrishoole (Irl), Bush (UK NI), Imsa and Drammen (Nor), Lagan (Swe), Midfjardara (Ice) | 0.98 | Dn |
|  | 1SW return to homewaters | 5 | Shannon, Screebe, Delphi, Bunowen and Burrishoole (Irl), Bush (UK NI), Imsa and Drammen (Nor), Lagan (Swe), Midfjardara (Ice) | 0.74 | Nt |
|  | 2SW return to homewaters | 10 | Midfjardara (Ice), Imsa and Drammen (Nor), Lagan (Swe) | 0.99 | Dn |
|  | 2SW return to homewaters | 5 | Midfjardara (Ice), Imsa and Drammen (Nor), Lagan (Swe) | 0.78 | Nt |
|  |  | Trends: | $\begin{aligned} & \mathrm{Up}=\text { significant increase } \\ & \mathrm{Dn}=\text { significant decrease } \\ & \mathrm{Nt}=\text { no trend } \end{aligned}$ |  |  |

Table 3.4.3.1 Wild adult counts to various rivers in the North East Atlantic area (Iceland, Sweden and Russia).

| Year | Iceland | Sweden | Russia | Russia | Russia | Russia | Russia | Russia | Russia | Russia | Russia |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | River <br> Ellidaar | River Högvadsån | River Ura | River Kitsa | River <br> Tuloma | River Varzuga | River Keret | $\begin{gathered} \text { River } \\ \text { Ponoy }{ }^{1} \\ \hline \end{gathered}$ | River Kola | River Yokanga | R. Zap. Litca |
|  | Estimate | Total trap | Total trap | $\begin{aligned} & \hline \text { Total } \\ & \text { Trap } \\ & \hline \end{aligned}$ | Total trap | Total trap | Total trap | $\begin{aligned} & \text { Total } \\ & \text { trap } \\ & \hline \end{aligned}$ | Total trap | Total trap | $\begin{gathered} \text { Total } \\ \text { trap } \\ \hline \end{gathered}$ |
| 1952 | 3,792 |  |  |  | 4,800 |  |  |  |  |  |  |
| 1953 | 2,526 |  |  |  | 2,950 |  |  |  |  |  |  |
| 1954 | 2,794 | 364 |  |  | 4,010 |  |  |  |  |  |  |
| 1955 | 4,118 | 210 |  |  | 4,600 |  |  |  | 4,855 |  |  |
| 1956 | 2,911 | 144 |  |  | 4,800 |  |  |  | 2,176 |  |  |
| 1957 | 2,965 | 126 |  |  | 4,300 |  |  |  | 2,949 |  |  |
| 1958 | 3,057 | 632 | 983 |  | 6,228 |  |  |  | 1,771 |  | 1,051 |
| 1959 | 4,773 | 197 | 997 |  | 6,125 |  |  |  | 2,790 |  | 1,642 |
| 1960 | 4,815 | 209 | 3,293 |  | 10,360 |  |  |  | 5,030 |  | 2,915 |
| 1961 | 3,779 | 229 | 2,178 |  | 11,050 | 55,480 |  |  | 5,121 |  | 2,091 |
| 1962 | 3,126 | 385 | 1,184 |  | 10,920 | 69,388 |  |  | 5,776 | 3,655 | 2,196 |
| 1963 | 4,031 | 217 | 811 |  | 7,880 | 64,210 |  |  | 3,656 | 3,253 | 1,983 |
| 1964 | 4,526 | 390 | 787 |  | 4,400 | 21,424 |  | 23,666 | 3,268 | 2,642 | 1,664 |
| 1965 | 3,249 | 442 | 1,334 |  | 5,600 | 63,812 |  | 12,998 | 3,676 | 4,482 | 1,506 |
| 1966 | 4,274 | 375 | 925 |  | 3,648 | 21,086 |  | 10,333 | 3,218 | 2,488 | 787 |
| 1967 | 4,839 | 90 | 2,679 |  | 9,011 | 20,534 |  | 11,527 | 7,170 | 4,993 | 1,486 |
| 1968 | 3,024 | 172 | 1,996 |  | 6,277 | 47,258 |  | 18,352 | 5,008 | 3,357 | 1,971 |
| 1969 | 3,580 | 321 | 967 |  | 4,538 | 53,048 |  | 9,267 | 6,525 | 1,437 | 2,341 |
| 1970 | 2,187 | 610 | 1,792 |  | 6,175 | 55,556 |  | 9,822 | 5,416 | 1,117 | 2,048 |
| 1971 | 2,590 | 173 | 1,172 |  | 3,284 | 71,400 |  | 8,523 | 4,784 | 2,300 | 1,502 |
| 1972 | 4,627 | 281 | 1,693 |  | 6,554 | 48,858 |  | 10,975 | 8,695 | 1,620 | 1,316 |
| 1973 | 6,014 | 100 | 2,502 | 4,472 | 9,726 | 45,750 |  | 20,553 | 9,780 | 869 | 1,319 |
| 1974 | 6,925 | 270 | 1,968 | 3,564 | 12,784 | 39,360 |  | 24,652 | 15,419 | 280 | 2,605 |
| 1975 | 7,184 | 138 | 3,249 | 13,950 | 11,074 | 89,836 |  | 41,666 | 12,793 | 736 | 2,456 |
| 1976 | 3,331 | 65 | 2,110 | 6,996 | 8,060 | 57,246 |  | 44,283 | 9,360 | 2,767 | 1,325 |
| 1977 | 3,756 | 49 | 2,784 | 7,976 | 2,878 | 35,354 |  | 37,159 | 7,180 | 2,488 | 1,595 |
| 1978 | 4,372 | 23 | 1,358 | 4,410 | 3,742 | 18,483 |  | 24,045 | 5,525 | 1,715 | 766 |
| 1979 | 4,948 | 15 | 888 | 5,998 | 2,887 | 40,992 |  | 17,920 | 6,281 | 598 | 700 |
| 1980 | 2,632 | 260 | 957 | 2,310 | 4,087 | 43,664 |  | 15,069 | 7,265 | 1,052 | 548 |
| 1981 | 2,656 | 512 | 438 | 5,013 | 3,467 | 32,158 |  | 11,670 | 7,131 | 472 | 477 |
| 1982 | 4,275 | 572 | 1,205 | 4,158 | 4,252 | 26,824 |  | 9,585 | 5,898 | 1,200 | 889 |
| 1983 | 3,257 | 447 | 2,108 | 3,778 | 9,102 | 59,784 |  | 15,594 | 10,643 | 1,769 | 1,254 |
| 1984 | 1,659 | 629 | 4,458 | 7,498 | 10,971 | 39,636 |  | 26,330 | 10,970 | 2,498 | 1,859 |
| 1985 | 2,896 | 768 | 2,634 | 11,134 | 8,067 | 48,566 |  | 38,787 | 6,163 | 1,774 | 1,563 |
| 1986 | 2,651 | 1,632 | 2,474 | 7,290 | 7,275 | 71,562 | 3,230 | 32,266 | 6,508 | 3,212 | 1,815 |
| 1987 | 2,191 | 1,475 | 1,788 | 9,911 | 5,470 | 137,419 | 3,427 | 21,212 | 6,300 | 3,468 | 1,498 |
| 1988 | 4,435 | 1,283 | 1,252 | 10,488 | 8,069 | 72,528 | 3,294 | 20,620 | 5,203 | 2,270 | 575 |
| 1989 | 4,329 | 480 | 2,434 | 3,697 | 8,413 | 65,524 | 3,531 | 19,214 | 10,929 | 2,850 | 2,613 |
| 1990 | 3,383 | 879 | 1,558 | 6,548 | 11,594 | 56,000 | 2,520 | 37,712 | 13,383 | 3,376 | 1,194 |
| 1991 | 3,020 | 534 | 1,328 | 3,041 | 7,253 | 63,000 | 690 | 21,000 | 8,500 | 1,704 | 2,081 |
| 1992 | 2,917 | 345 | 3,391 | 8,587 | 5,377 | 61,300 | 536 | 26,600 | 14,670 | 5,208 | 2,755 |
| 1993 | 3,363 | 603 | 1,972 | 2,956 | 4,516 | 68,300 | 687 | 26,800 | 11,400 | 2,600 | 2,267 |
| 1994 | 2,298 | 640 | 1,738 | 3,222 | 3,316 | 77,800 | 753 | 28,600 | 9,730 | 2,500 | 2,100 |
| 1995 | 2,509 | 156 | 1,461 | 3,207 | 4,737 | 42,290 | 1,066 | 33,100 | 6,051 | 1,153 | 1,916 |
| 1996 | 2,170 | 249 | 1,171 | 4,740 | 4,424 | 67,900 | 391 | 32,600 | 7,700 | 2,700 | 2,330 |
| 1997 | 1,132 | 189 | 2,028 | 5,222 | 4,405 | 73,430 | 180 | 37,600 | 6,180 | 2,700 | 1,350 |
| 1998 | 875 | 160 | 1,100 | 5,560 | 3,338 | 83,050 | 607 | 34,400 | 4,848 | - | 1,510 |
| 1999 | 628 | 450 | 2,180 | 4,300 | 6,040 | 71,000 | 333 | 20,300 | 7,950 | - | 1,720 |
| 2000 | 1,113 | 653 | 780 | 4,170 | 6,660 | 75,540 | 974 | 23,000 | 6,360 | - | 910 |
| Mean |  |  |  |  |  |  |  |  |  |  |  |
| 95-99 | 1,463 | 241 | 1,588 | 4,606 | 4,589 | 67,534 | 515 | 31,600 | 6,546 | 2,184 | 1,765 |

[^6]Continued...

| Year | Russia | UK (E\&W) | UK (E\&W) | UK (E\&W) | UK (E\&W) | UK (E\&W) | UK (E\&W) | UK (E\&W) | UK (E\&W) | UK (E\&W) | UK (NI) | UK (NI) | UK (NI) | UK (NI) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Umba | Frome | Test | Itchen | Kent | Leven | Tamar | Dee | Lune | Caldew | Roe | Bush | Faughan | Mourne |
|  | Total trap | Counter | $\begin{aligned} & \text { Counter } \\ & + \text { catch } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { Counter } \\ & + \text { catch } \\ & \hline \end{aligned}$ | Counter | Counter | Counter | $\begin{aligned} & \hline \text { Counter } \\ & + \text { catch } \\ & \hline \end{aligned}$ | $\begin{aligned} & \hline \text { Counter } \\ & + \text { catch } \\ & \hline \end{aligned}$ | Trap | Counter | Total trap | Counter | Counter |
| 1966 |  |  |  |  |  |  |  |  |  |  |  |  | 6,792 | 15,112 |
| 1967 |  |  |  |  |  |  |  |  |  |  |  |  | 1,723 | 7,087 |
| 1968 |  |  |  |  |  |  |  |  |  |  |  |  | 1,657 | 2,147 |
| 1969 | 2,030 |  |  |  |  |  |  |  |  |  |  |  | 1,195 | 1,569 |
| 1970 | 1,316 |  |  |  |  |  |  |  |  |  |  |  | 3,214 | 5,050 |
| 1971 | 288 |  |  |  |  |  |  |  |  |  |  |  | 1,758 | 4,401 |
| 1972 | 548 |  |  |  |  |  |  |  |  |  |  |  | 1,020 | 1,453 |
| 1973 | 2,536 |  |  |  |  |  |  |  |  |  |  | 2,614 | 1,885 | 2,959 |
| 1974 | 2,692 |  |  |  |  |  |  |  |  |  |  | 3,483 | 2,709 | 3,630 |
| 1975 | 5,432 |  |  |  |  |  |  |  |  |  |  | 3,366 | 1,617 | 1,742 |
| 1976 | 1,926 |  |  |  |  |  |  |  |  |  |  | 3,124 | 2,040 | 2,259 |
| 1977 | 3,692 |  |  |  |  |  |  |  |  |  |  | 1,775 | 2,625 | 2,419 |
| 1978 | 3,308 |  |  |  |  |  |  |  |  |  |  | 1,621 | 2,587 | 5,057 |
| 1979 | 3,772 |  |  |  |  |  |  |  |  |  |  | 1,820 | 3,262 | 2,226 |
| 1980 | 5,924 |  |  |  |  |  |  |  |  |  |  | 2,863 | 3,288 | 3,146 |
| 1981 | 6,252 |  |  |  |  |  |  |  |  |  |  | 1,539 | 3,772 | 2,399 |
| 1982 | 8,690 |  |  |  |  |  |  |  |  |  |  | 1,571 | 2,909 | 4,755 |
| 1983 | 7,850 |  |  |  |  |  |  |  |  |  |  | 1,030 | 2,410 | 1,271 |
| 1984 | 6,326 |  |  |  |  |  |  |  |  |  |  | 6,721 | 2,116 | 1,877 |
| 1985 | 12,190 |  |  |  |  |  |  |  |  |  |  | 2,443 | 9,077 | 8,149 |
| 1986 | 8,568 |  |  |  |  |  |  |  |  |  |  | 2,930 | 4,915 | 6,295 |
| 1987 | 10,040 |  |  |  |  |  |  |  |  |  |  | 2,530 | 907 | 2,322 |
| 1988 | 8,455 | 4,093 | 1,507 | 1,336 |  |  |  |  |  |  |  | 2,832 | 3,228 | 7,572 |
| 1989 | 12,029 | 3,186 | 1,730 | 791 | 1,137 |  |  |  | 8,785 |  |  | 1,029 | 8,287 | 9,497 |
| 1990 | 9,040 | 1,880 | 790 | 367 | 2,216 |  |  |  | 8,261 |  |  | 1,850 | 6,458 | 11,541 |
| 1991 | 6,400 | 805 | 538 | 152 | 1,736 |  |  |  | 7,591 |  |  | 2,341 | 4,301 | 7,987 |
| 1992 | 8,400 | 900 | 614 | 357 | 1,816 |  |  | 4,643 | 4,066 |  |  | 2,546 | 7,375 | 7,420 |
| 1993 | 8,500 | 1,182 | 1,155 | 852 | 1,526 | 101 |  | 9,757 | 7,883 |  |  | 3,235 | 8,655 | 17,855 |
| 1994 | 6,800 | 1,078 | 775 | 375 | 2,072 | 102 | 6,359 | 8,285 | 6,254 | 1,590 |  | 2,010 | 7,439 | 19,908 |
| 1995 | 7,340 | 1,016 | 647 | 880 | 1,396 | 123 | 5,637 | 5,703 | 4,589 | 1,417 |  | 1,521 | 5,838 | 7,547 |
| 1996 | 6,450 | 1,353 | 623 | 433 | 1,219 | 155 | 3,988 | 4,931 | 4,739 | 1,289 |  | 1,097 | 13,297 | 5,475 |
| 1997 | 6,200 | 1,157 | 361 | 246 | 491 | 41 | 2,989 | 5,495 | 3,121 | 889 |  | 1,677 | 4,500 | 6,979 |
| 1998 | 6,440 | 1,210 | 898 | 453 | 800 | 39 | 4,176 | 6,661 | 7,457 | 1,106 | 2,600 | 2,995 | $\mathrm{n} / \mathrm{a}$ | 6,077 |
| 1999 | 6,850 | $\mathrm{n} / \mathrm{a}$ | 867 | 213 | 1,018 | 98 | 3,588 | 3,664 | 4,936 | 1,022 | n/a | 959 | n/a | 8,500 |
| 2000 | 6,780 | $\mathrm{n} / \mathrm{a}$ | 583 | 208 | 2,354 | 322 | 3,539 | 3,751 | 8,383 | 1,566 | 3,643 | 950 | 2,551 | 4,690 |
| $\begin{gathered} \text { Mean } \\ 95-99 \\ \hline \end{gathered}$ | 6,656 | 1,184 | 679 | 445 | 985 | 91 | 4,076 | 5,291 | 4,968 | 1,145 | - | 1,650 | 7,878 | 6,916 |


| Table 3.4.3.1 Cont'd. | Wild adult counts to various rivers in the North East Atlantic area (UK, France, Norway and Ireland). |  |  |  |  |  |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| UK(Scotl.) | UK(Scotl.) | UK(Scotl.) | UK(Scotl.) | France | France | France France Norway |


|  | UK(Scotl.) | UK(Scotl.) | UK(Scotl.) | UK(Scotl.) | France | France | France | France | Norway | Norway | Norway | Ireland |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | N. Esk | West Water | Girnock | Baddoch | Nivelle | Oir | Scorff | Bresle | Halselva | Imsa | Orkla | Burrishoole |
|  | Counter | Counter | Total trap Females | Total trap Females | Trap est. | Trap est. | Trap est. | Trap est. | Total trap | Total trap | Counter | Total trap |
| 1966 |  |  | 156 |  |  |  |  |  |  |  |  |  |
| 1967 |  |  | 115 |  |  |  |  |  |  |  |  |  |
| 1968 |  |  | 111 |  |  |  |  |  |  |  |  |  |
| 1969 |  |  | 31 |  |  |  |  |  |  |  |  |  |
| 1970 |  |  | 34 |  |  |  |  |  |  |  |  |  |
| 1971 |  |  | 61 |  |  |  |  |  |  |  |  |  |
| 1972 |  |  | 79 |  |  |  |  |  |  |  |  |  |
| 1973 |  |  | 127 |  |  |  |  |  |  |  |  |  |
| 1974 |  |  | 105 |  |  |  |  |  |  |  |  |  |
| 1975 |  |  | 65 |  |  |  |  |  |  |  |  |  |
| 1976 |  |  | 90 |  |  |  |  |  |  |  |  |  |
| 1977 |  |  | 49 |  |  |  |  |  |  |  |  |  |
| 1978 |  |  | 16 |  |  |  |  |  |  |  |  |  |
| 1979 |  |  | 49 |  |  |  |  |  |  |  |  |  |
| 1980 |  |  | 121 |  |  |  |  |  |  |  |  | 832 |
| 1981 | 9,025 |  | 41 |  |  |  |  |  |  |  |  | 348 |
| 1982 | 8,121 |  | 43 |  |  |  |  |  |  | 66 |  | 510 |
| 1983 | 8,972 |  | 26 |  |  |  |  |  |  | 14 |  | 602 |
| 1984 | 7,007 |  | 58 |  | 33 | 307 |  | 110 |  | 32 |  | 319 |
| 1985 | 9,912 |  | 30 |  | 61 | 296 |  | 135 |  | 31 |  | 567 |
| 1986 | 6,987 |  | 75 |  | 204 | 216 |  | 210 |  | 22 |  | 495 |
| 1987 | 7,014 |  | 110 |  | 138 | 180 |  | 200 | 52 | 9 |  | 468 |
| 1988 | 11,243 |  | 112 | 47 | 130 | 235 |  | 105 | 77 | 44 |  | 458 |
| 1989 | 11,026 |  | 43 | 67 | 263 | 235 |  | 220 | 64 | 83 |  | 662 |
| 1990 | 4,762 |  | 29 | 52 | 291 | 84 |  | 125 | 68 | 67 |  | 231 |
| 1991 | 9,127 | 2,962 | 57 | 46 | 184 | 47 |  | 215 | 89 | 43 |  | 547 |
| 1992 | 10,795 | 2,809 | 35 | 32 | 234 | 60 |  | 225 | 35 | 70 |  | 360 |
| 1993 | 10,887 | 2,699 | 21 | 27 | 472 | 176 |  | 75 | 18 | 39 |  | 528 |
| 1994 | 11,341 | 2,976 | 37 | 40 | 317 | 155 | $694{ }^{1}$ | 105 | 29 | 30 | 4,305 | 516 |
| 1995 | 9,864 | 2,391 | 71 | 16 | 195 | 128 | 982 | 80 | 9 | 1 | - | 561 |
| 1996 | 7,993 | 2,656 | 41 | 26 | 214 | 196 | 756 | 40 | 25 | 2 | 4,405 | 405 |
| 1997 | 11,315 | 2,926 | 9 | 9 | 126 | 67 | 542 | 45 | 77 | 9 | 2,509 | 538 |
| 1998 | 10,474 | 2,422 | 11 | 10 | 160 | 189 | 551 | 270 | 38 | 20 | 4,171 | 516 |
| 1999 | 11,789 | 2,312 | 22 | 22 | 160 | 257 | 353 | 62 | 14 | 36 | 2,827 | 508 |
| 2000 | 8,353 | 2,092 | 27 | 9 | 151 | 490 | n/a | 35 | 25 | 8 | 7,719 | 574 |
| Mean |  |  |  |  |  |  |  |  |  |  |  |  |
| 95-99 | 10,287 | 2,541 | 31 | 17 | 171 | 167 | 637 | 99 | 33 | 14 | 3,478 | 506 |

Table 3.4.4.1 Estimated survival of wild smolts (\%) to return to homewaters (prior to coastal fisheries) for various monitored rivers in the NE Atlantic area.

| Smolt migration year | Iceland ${ }^{1}$ |  |  |  |  | Ireland |  | $\begin{gathered} \text { UK (N.Ireland) }{ }^{8} \\ \hline \text { R. Bush } \\ 1 \mathrm{SW}^{3} \\ \hline \end{gathered}$ | Norway $^{2}$ |  | $\text { UK (Scotland) }{ }^{2}$ |  |  | France |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ellidaar 1SW | $\begin{aligned} & \hline \text { R.Vest } \\ & \text { 1SW } \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { dalsa }^{4} \\ & 2 \mathrm{SW} \\ & \hline \end{aligned}$ | R.Midfjardara ${ }^{4}$ |  | $\begin{gathered} \text { River Corrib } \\ \text { 1SW } \\ \hline \end{gathered}$ | $\begin{gathered} \text { River Corrib } \\ 2 \mathrm{SW} \\ \hline \end{gathered}$ |  | R. Imsa 1SW | 2SW | North Esk |  |  | Nivelle ${ }^{6}$ <br> All ages | $\begin{gathered} \text { Bresle } \\ \text { All ages } \\ \hline \end{gathered}$ |
| 1975 | 20.8 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1980 |  |  |  |  |  | 14.3 | 0.5 |  |  |  |  |  |  |  |  |
| 1981 |  |  |  |  |  | 10.0 | 3.0 |  | 17.3 | 4.0 | 13.7 | 6.9 | 0.3 |  |  |
| 1982 |  |  |  |  |  | 16.7 | 2.7 |  | 5.3 | 1.2 | 12.6 | 5.4 | 0.2 |  |  |
| 1983 |  | 2.0 |  |  |  | 8.0 | 1.5 |  | 13.5 | 1.3 | - | - | - |  |  |
| 1984 |  |  |  |  |  | 20.0 | 1.6 |  | 12.1 | 1.8 | 10.0 | 4.1 | 0.1 |  |  |
| 1985 | 9.4 |  |  |  |  | 15.2 | 1.4 |  | 10.2 | 2.1 | 26.1 | 6.4 | 0.2 |  |  |
| 1986 |  |  |  |  |  | - | - | 31.3 | 3.8 | 4.2 | - | - | - | 15.1 |  |
| 1987 |  |  |  | 2.4 | 1.4 | 13.3 | 0.6 | 35.1 | 17.3 | 5.6 | 13.9 | 3.4 | 0.1 | 2.6 |  |
| 1988 | 12.7 |  |  | 0.6 | 0.9 | 11.7 | 0.6 | 36.2 | 13.3 | 1.1 | - | - | - | 2.4 |  |
| 1989 | 8.1 | 1.1 | 2.0 | 0.2 | 0.7 | 5.3 | 0.6 | 25.0 | 8.7 | 2.2 | 7.8 | 4.9 | 0.1 | 3.5 |  |
| 1990 | 5.4 | 1.0 | 1.0 | 1.2 | 1.3 | 4.0 | 0.5 | 34.7 | 3.0 | 1.3 | 7.3 | 3.1 | 0.2 | 1.8 |  |
| 1991 | 8.8 | 4.2 | 0.6 | 1.1 | 0.5 | 5.8 | 1.0 | 27.8 | 8.7 | 1.2 | 11.2 | 4.5 | - | 9.2 |  |
| 1992 | 9.6 | 2.4 | 0.8 | 1.4 | 0.5 | 5.8 | - | 29.0 | 6.7 | 0.9 | - | - | - | 8.9 | $6.9{ }^{7}$ |
| 1993 | 9.8 | - | - | 1.0 | 1.1 | 8.7 | 1.6 | - | 15.6 |  | - | - | - | $8.3{ }^{7}$ | $10.3{ }^{7}$ |
| 1994 | 9.0 | - | - | 1.4 | 0.6 | 7.8 | 1.1 | 27.1 | - | - | 17.2 | 2.3 | 0.1 | $7.2{ }^{7}$ | $7.5{ }^{7}$ |
| 1995 | 9.4 | 1.6 | 1.2 | 0.3 | 0.9 | 6.7 | 0.1 | n/a | 1.8 | 1.5 | 11.5 | 5.1 | 0.1 | 2.3 | - |
| 1996 | 4.6 | 1.4 | 0.3 | 1.2 | 0.7 | 5.1 | 0.9 | 31.0 | 3.5 | 0.9 | 10.7 | 3.5 | 0.2 | 4.4 | - |
| 1997 | 5.3 | 0.7 | 0.5 | 2.4 | 0.5 | 10.2 | 0.6 | 19.8 | 1.5 | 0.3 | 10.3 | 6.3 | 0.1 | 3.4 | 4.8 |
| 1998 | 5.3 | 1.0 | 1.0 | 1.3 | - | 4.4 | 0.8 | 13.4 | 7.2 | 1.1 | - | - |  | 2.6 | - |
| 1999 | 7.7 | 1.3 |  | - |  | 4.3 |  | 16.5 | 3.7 |  | - |  |  | - | - |
| $\begin{aligned} & \text { Mean } \\ & \text { (5-year) } \end{aligned}$ | 6.7 | 1.2 | 0.7 | 1.3 | 0.8 | 6.8 | 0.9 | 22.8 | 3.5 | 0.9 | 12.4 | 4.3 | 0.1 | 3.2 | 4.8 |
| (10-year) | 7.5 | 1.7 | 0.9 | 1.2 | 0.8 | 6.4 | 0.8 | 26.0 | 6.3 | 1.2 | 10.9 | 4.2 | 0.1 | 4.5 | 4.8 |

${ }^{1}$ Microtags.
${ }^{2}$ Carlin tags, not corrected for tagging mortality.
${ }^{3}$ Microtags, corrected for tagging mortality.
${ }^{4}$ Assumes $50 \%$ exploitation in rod fishery.
${ }^{5}$ Minimum estimates.
${ }^{6}$ From $0+$ stage in autumn.
${ }^{7}$ Incomplete returns.
${ }^{8}$ Assumes $30 \%$ exploitation in trap fishery.

Table 3.4.4.2 Estimated survival of hatchery smolts (\%) to adult return to homewaters, (prior to coastal fisheries) for monitored rivers and experimental facilities in the NE Atlantic area.

| Smolt year | Iceland $^{1}$ |  | UK (N. Ireland) ${ }^{1}$ |  | Norway ${ }^{2}$ |  |  |  | Sweden ${ }^{2}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | R Midfiardara ${ }^{3}$ |  | R. Bush (1SW) |  | R. Imsa |  | R. Drammen |  | R. Lagan |  |
|  | 1SW | 2SW | $1+$ smolts | $2+$ smolts | 1SW | 2SW | 1SW | 2SW | 1SW | 2SW |
| 1981 |  |  |  |  | 10.1 | 1.3 |  |  |  |  |
| 1982 |  |  |  |  | 4.2 | 0.6 |  |  |  |  |
| 1983 | 0.0 | 0.2 | 1.9 | 8.1 | 1.6 | 0.1 |  |  |  |  |
| 1984 | 0.5 | 0.2 | 13.3 |  | 3.8 | 0.4 | 3.5 | 3.0 | 11.8 | 1.1 |
| 1985 | 0.4 | 0.1 | 15.4 | 17.5 | 5.8 | 1.3 | 3.4 | 1.9 | 11.8 | 0.9 |
| 1986 | 0.4 | 0.7 | 2.0 | 9.7 | 4.7 | 0.8 | 6.1 | 2.2 | 7.9 | 2.5 |
| 1987 | 2.7 | 0.7 | 6.5 | 19.4 | 9.8 | 1.0 | 1.7 | 0.7 | 8.4 | 2.4 |
| 1988 | 0.7 | 0.2 | 4.9 | 6.0 | 9.5 | 0.7 | 0.5 | 0.3 | 4.3 | 0.6 |
| 1989 | 0.7 | 0.4 | 8.1 | 23.2 | 3.0 | 0.9 | 1.9 | 1.3 | 5.0 | 1.3 |
| 1990 | 1.9 | 0.5 | 5.6 | 5.6 | 2.8 | 1.5 | 0.3 | 0.4 | 5.2 | 3.1 |
| 1991 | 1.8 | 0.2 | 5.4 | 8.8 | 3.2 | 0.7 | 0.1 | 0.1 | 3.6 | 1.1 |
| 1992 | 1.3 | 0.2 | 6.0 | 7.8 | 3.8 | 0.7 | 0.4 | 0.6 | 1.5 | 0.4 |
| 1993 | 0.5 | 0.2 | 1.1 | 5.8 | 6.5 | 0.5 | 3.0 | 1.0 | 2.6 | 0.9 |
| 1994 | 1.0 | 0.2 | 1.6 | - | 6.2 | 0.6 | 1.2 | 0.9 | 4.0 | 1.2 |
| 1995 | 0.8 | 0.1 | 3.1 | 2.4 | 0.4 | 0.0 | 0.7 | 0.3 | 3.9 | 0.6 |
| 1996 | 0.1 | 0.0 | 2.0 | 2.3 | 2.1 | 0.2 | 0.3 | 0.2 | 3.5 | 0.5 |
| 1997 | 0.9 | 0.0 | no release | 4.1 | 0.9 | 0.0 | 0.4 | 0.1 | 0.6 | 0.5 |
| 1998 | no release | no release | 2.3 | 4.5 | 2.4 | 0.1 | 1.9 | 0.6 | 1.6 | 0.9 |
| 1999 | no release |  | 2.7 | 5.8 | 6.0 |  | 1.7 |  | 2.1 |  |
| Mean |  |  |  |  |  |  |  |  |  |  |
| (5-year) | 1.1 | 0.3 | 4.1 | 8.3 | 4.3 | 0.7 | 1.4 | 0.8 | 4.0 | 1.2 |
| (10-year) | 1.0 | 0.2 | 3.9 | 7.2 | 3.1 | 0.6 | 1.0 | 0.5 | 3.2 | 1.0 |

${ }^{1}$ Microtagged.
${ }^{2}$ Carlin tagged, not corrected for tagging mortality.

Table 3.4.4.2 Cont'd. Estimated survival of hatchery smolts (\%) to 1SW adult return to homewaters, (prior to coastal fisheries) for monitored rivers and experimental facilities in Ireland.

| Smolt year | R. Shannon | R. Screebe | R. Burrishoole $^{1}$ | R. Delphi | R. <br> Bunowen | R. Lee | R. Corrib Cong. 2 | R. Corrib Galway 2 | R. Erne |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 8.6 |  |  |  |  | 10.8 | 0.9 |  |  |
| 1981 | 2.8 |  | 9.1 |  |  | 2 | 1.2 |  |  |
| 1982 | 4.1 |  | 9.9 |  |  | 16.3 | 2.7 | 16.1 |  |
| 1983 | 3.9 |  | 4.3 |  |  | 2 | 1.7 | 4.1 |  |
| 1984 | 4.9 | 10.4 | 26.9 |  |  | 0.1 | 5.2 | 13.2 | 9.3 |
| 1985 | 4.8 | 12.3 | 27.9 |  |  | 17.7 | 1.4 | 14.4 | 9.9 |
| 1986 | 9.1 | 0.4 | 8.8 |  |  | 16.3 | - | 7.6 | 10.1 |
| 1987 | 4.7 | 8.3 | 13.8 |  |  | 8.6 | - | 2.2 | 6.9 |
| 1988 | 4.9 | 9.2 | 17.1 |  |  | 5.5 | 4.2 | - | 2.6 |
| 1989 | 5.0 | 1.6 | 10.1 |  |  | 1.7 | 6 | 4.9 | 1.2 |
| 1990 | 1.3 | 0.0 | 12.1 |  |  | 2.5 | 0.2 | 2.3 | 1.3 |
| 1991 | 4.1 | 0.2 | 12.8 | 10.8 |  | 0.8 | 3.5 | 4 | 1.3 |
| 1992 | 4.3 | 1.3 | 7.1 | 10.0 | 4.2 | - | 0.9 | 0.6 | - |
| 1993 | 2.9 | 2.2 | 14.0 | 14.3 | 5.4 | - | 1 | - | - |
| 1994 | 5.1 | 1.9 | 13.1 | 7.6 | 8.1 | - | - | 5.3 | - |
| 1995 | 3.6 | 4.1 | 8.5 | 2.5 | 3.4 | - | 2.4 | - | - |
| 1996 | 2.9 | 1.8 | 5.5 | 9.9 | 3.3 | - | - | - | - |
| 1997 | 6.0 | 0.4 | 13.3 | 10.8 | 5.1 | 6.9 | - | - | 5.9 |
| 1998 | 2.8 | 1.3 | 4.9 | 5.9 | 2.6 | 4.6 | 3.3 | 2.9 | 1.9 |
| 1999 | 0.5 | 2.5 | 8.6 | 7.8 | 1.4 | - | - | 3.2 | 3.5 |
| Mean |  |  |  |  |  |  |  |  |  |
| (5-year) | 4.1 | 1.9 | 9.1 | 7.3 | 4.5 | 5.8 | 2.9 | 4.1 | 3.9 |
| (10-year) | 3.8 | 1.5 | 10.1 | 9.0 | 4.6 | 3.3 | 2.5 | 3.3 | 2.3 |

[^7]Table 3.5.1 Assessment of the effects of the suspension of commercial fishing at Faroes on the numbers of salmon returning to home waters.

Fishing season

|  | 1991/92 | 1992/93 | 1993/94 | 1994/95 | 1995/96 | 1996/97 | 1997/98 | 1998/99 | 1999/00 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| NASCO quota ( t ) for the calender year if fishery operated ${ }^{\text {a }}$ | 550 | 550 | 550 | 550 | 470 | 425 | 380 | 330 | 300 |
| Expected No. fish landed if quota had been taken ${ }^{\text {b }}$ | 147,048 | 162,850 | 182,027 | 172,931 | 142,037 | 128,438 | 140,927 | 122,384 | 111,258 |
| Discard rate | 8.8\% | 9.4\% | 14.4\% | 15.1\% | $11.9 \%{ }^{\text {c }}$ | $11.9 \%{ }^{\text {c }}$ | 16.9\% | $16.9 \%{ }^{\text {g }}$ | $16.9 \%{ }^{\text {g }}$ |
| Discard mortality | 80.0\% | 80.0\% | 80.0\% | 80.0\% | 80.0\% | 80.0\% | 80\% | 80\% | 80\% |
| Expected No. fish killed if fishery operated | 158,399 | 176,367 | 206,524 | 197,536 | 157,422 | 142,350 | 163,855 | 142,295 | 129,359 |
| No. fish killed in research fishery | 9,350 | 9,099 | 3,035 | 4,187 | 282 | 0 | 1465 | 0 | $2967{ }^{\text {h }}$ |
| Total number of fish saved per year | 149,049 | 167,268 | 203,489 | 193,349 | 157,140 | 142,350 | 162,390 | 142,295 | 126,392 |
| Proportion of farmed fish in catch | 37.0\% | 27.0\% | 17.0\% | 19.0\% | 19.0\% | 19.0\% | 19.0\% | 19.0\% ${ }^{\text {f }}$ | $19.0 \%{ }^{\text {f }}$ |
| Number farm escapees spared | 55,148 | 45,162 | 34,593 | 36,736 | 29,857 | 27,046 | 30,854 | 27,036 | 24,015 |
| Number of wild fish spared | 93,901 | 122,106 | 168,896 | 156,613 | 127,283 | 115,303 | 131,536 | 115,259 | 102,378 |
| Sea age composition of wild fish: 1SW | 4.0\% | 12.0\% | 16.0\% | 10.6\% | $10.7 \%{ }^{\text {d }}$ | $10.7 \%{ }^{\text {d }}$ | 19.2\% | 19.2\% | 19.2\% |
| 2SW | 83.0\% | 61.0\% | 64.0\% | 80.8\% | $72.2 \%{ }^{\text {d }}$ | $72.2 \%{ }^{\text {d }}$ | 74.6\% | 74.6\% | 74.6\% |
| 2SW+ | 13.0\% | 27.0\% | 20.0\% | 8.6\% | $17.2 \%{ }^{\text {d }}$ | $17.2 \%{ }^{\text {d }}$ | 6.2\% | 6.2\% | 6.2\% |
|  | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| Additional salmon 1SW <br> expected to have returned: MSW | $\begin{array}{r} \mathbf{2 , 8 4 2} \\ \mathbf{7 0 , 8 0 9} \end{array}$ | $\begin{array}{r} 11,429 \\ 106,307 \end{array}$ | $\begin{array}{r} 21,078 \\ 134,159 \end{array}$ | $\begin{array}{r} 12,949 \\ 138,533 \end{array}$ | $\begin{array}{r} 10,573 \\ 122,196 \end{array}$ | $\begin{array}{r} 9,578 \\ 105,368 \end{array}$ | $\begin{array}{r} 19,699 \\ 103,169 \end{array}$ | $\begin{aligned} & \mathbf{1 7 , 2 6 1} \\ & \mathbf{9 9 , 1 3 0} \end{aligned}$ | $\begin{aligned} & \mathbf{1 5 , 3 3 2} \\ & \mathbf{8 7 , 7 2 6} \end{aligned}$ |
| Estimated 1SW returns to all Eurom homewaters: ${ }^{\text {e }}$ | 2,071,416 | 1,976,751 | 2,349,998 | 1,906,497 | 1,574,717 | 1,407,466 | 1752513 | 1241462 | 1661995 |
| \% 1SW returns derived from suspension of commerial fishing at Faroes: | 0\% | 1\% | 1\% | 1\% | 1\% | 1\% | 1\% | 1\% | 1\% |
| Estimated MSW returns to all European homewaters: ${ }^{\text {e }}$ | 990,822 | 1,002,492 | 1,134,800 | 1,034,176 | 953,108 | 708,513 | 755788 | 751153 | 919223 |
| \% MSW returns derived from suspension of commerial fishing at Faroes: | 7\% | 11\% | 12\% | 13\% | 13\% | 15\% | 14\% | 13\% | 10\% |
| Estimated 1SW returns to Northern European homewaters: ${ }^{\text {e }}$ | 718,658 | 660,583 | 851,641 | 596,837 | 512,841 | 551,205 | 668824 | 614962 | 842672 |
| $\% 1 \mathrm{SW}$ returns derived from suspension of commerial fishing at Faroes: (Assuming 65\% from N. Europe) | 0\% | 1\% | 2\% | 1\% | 1\% | 1\% | 2\% | 2\% | 1\% |
| Estimated MSW returns to Momewaters: ${ }^{\text {e }}$ | 457,004 | 506,992 | 489,097 | 445,638 | 484,430 | 365,885 | 396970 | 443141 | 582598 |
| \% MSW returns derived from suspension of commerial fishing at Faroes: (Assuming 65\% from N. Europe) | 10\% | 14\% | 18\% | 20\% | 16\% | 19\% | 17\% | 15\% | 10\% |

a. NASCO quota agreed for the calender year in the latter part of the fishing season.
b. Expected no. landed in year $y$ calculated from quota: $\operatorname{Sum}\left(\mathrm{p}_{i} / \mathrm{w}_{i}\right) *$ Quota ${ }_{v}, \mathrm{p}_{i}$ is proportion of age group $i, i=1 ; 2$ and $2+\mathrm{SW}$, and $\mathrm{w}_{i}$ is mean weight of sea age $i$.
c. No data, estimated from mean discard rate 1992-95.
d. No data, mean values from 1992-95 data.
e. Includes farmed escapees.
f. Data not yet available, mean value from 1994-1996 data
g. Taken from 1997/98 research fishery
h. In 1999/2000 a commercial fishery operated instead of a research fishery

Table 3.5.2. Reduction in gear units over the period 1991-2000 for countries where such information is available.

| Country | Type of gear units | \% Change in gear units <br> over 1991 to 2000 |
| :--- | :--- | :---: |
| UK (England \& Wales) | Gillnet <br> Sweepnet <br> Hand-held net <br> Fixed engine | -45 |
|  | Fixed engine <br> Net and coble | -46 |
| UK (Scotland) | Driftnet <br> Draftnet <br> Bagnets and boxes | -58 |
| UK (N. Ireland) | Bagnet <br> Bendnet | -84 |
|  | Norway | Driftnet <br> Draftnet <br> Other nets |
| Ireland | Commercial nets in freshwater | -74 |
| France | -44 |  |

a Information on the number of gear units deployed in France for 2000 are not available. The \% change in gear units presented covers the period 1991-1999.

Table 3.6.2.1 Estimated number of RETURNING 1 SW salmon by NEAC country and year

| Year | Northern Europe |  |  |  |  |  |  | Southern Europe |  |  |  |  |  |  | NEAC Area |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Finland | Iceland | Norway | Russia | Sweden | Total |  | France | Ireland | UK(EW) | UK(NI) | UK(Scot) | Total |  | Total |  |
|  |  |  |  |  |  | Est. | SD |  |  |  |  |  | Est. | SD | Est. | SD |
| 1971 | 7,769 | 63,237 | 548,639 | 123,242 | 11,868 | 754,754 | 76,957 | 53,427 | 1,070,447 | 167,317 | 186,420 | 1,168,002 | 2,645,613 | 203,843 | 3,400,367 | 217,886 |
| 1972 | 12,058 | 64,211 | 706,255 | 135,241 | 9,546 | 927,311 | 98,959 | 105,359 | 1,164,565 | 186,085 | 161,965 | 1,073,691 | 2,691,664 | 192,790 | 3,618,976 | 216,704 |
| 1973 | 18,278 | 66,016 | 783,174 | 227,366 | 11,712 | 1,106,546 | 109,210 | 64,211 | 1,256,423 | 177,264 | 144,386 | 1,260,818 | 2,903,101 | 212,604 | 4,009,648 | 239,013 |
| 1974 | 16,906 | 46,644 | 736,277 | 207,967 | 17,307 | 1,025,101 | 105,314 | 30,123 | 1,379,752 | 174,495 | 156,802 | 1,317,677 | 3,058,849 | 232,899 | 4,083,950 | 255,603 |
| 1975 | 16,819 | 63,731 | 686,855 | 222,157 | 18,386 | 1,007,948 | 98,045 | 60,740 | 1,429,273 | 200,739 | 127,179 | 1,080,391 | 2,898,323 | 212,167 | 3,906,270 | 233,726 |
| 1976 | 14,490 | 53,341 | 684,807 | 167,519 | 10,434 | 930,591 | 98,129 | 55,623 | 1,007,550 | 125,363 | 87,787 | 935,194 | 2,211,517 | 180,113 | 3,142,107 | 205,109 |
| 1977 | 13,094 | 63,767 | 671,580 | 141,326 | 4,888 | 894,654 | 96,686 | 43,066 | 890,183 | 143,303 | 86,806 | 1,011,664 | 2,175,022 | 184,478 | 3,069,676 | 208,279 |
| 1978 | 9,552 | 77,629 | 724,892 | 153,472 | 5,766 | 971,310 | 101,399 | 43,785 | 795,426 | 141,288 | 112,183 | 1,102,254 | 2,194,935 | 198,692 | 3,166,245 | 223,070 |
| 1979 | 9,876 | 68,943 | 821,339 | 175,948 | 5,927 | 1,082,032 | 115,874 | 49,724 | 709,136 | 109,888 | 79,918 | 947,027 | 1,895,693 | 159,198 | 2,977,725 | 196,903 |
| 1980 | 9,737 | 35,260 | 824,362 | 115,940 | 7,378 | 992,677 | 114,246 | 104,147 | 543,181 | 139,614 | 102,394 | 589,955 | 1,479,291 | 107,608 | 2,471,968 | 156,945 |
| 1981 | 9,008 | 49,585 | 569,815 | 83,525 | 13,250 | 725,184 | 78,770 | 82,959 | 373,334 | 170,141 | 80,557 | 738,379 | 1,445,370 | 118,351 | 2,170,554 | 142,168 |
| 1982 | 6,446 | 32,731 | 415,943 | 127,541 | 11,852 | 594,513 | 58,571 | 51,006 | 669,910 | 101,354 | 115,640 | 1,029,799 | 1,967,709 | 170,176 | 2,562,222 | 179,973 |
| 1983 | 9,765 | 51,748 | 690,376 | 185,222 | 15,598 | 952,710 | 61,855 | 55,263 | 1,100,043 | 137,806 | 162,074 | 1,105,092 | 2,560,278 | 193,541 | 3,512,988 | 203,185 |
| 1984 | 12,223 | 32,817 | 872,447 | 175,950 | 22,290 | 1,115,727 | 76,051 | 90,802 | 524,866 | 117,621 | 63,674 | 1,130,608 | 1,927,570 | 182,348 | 3,043,298 | 197,572 |
| 1985 | 15,411 | 57,435 | 916,774 | 231,986 | 26,626 | 1,248,231 | 83,554 | 33,478 | 1,052,958 | 124,252 | 82,531 | 852,558 | 2,145,776 | 158,062 | 3,394,008 | 178,787 |
| 1986 | 15,369 | 83,403 | 767,763 | 203,397 | 28,737 | 1,098,669 | 70,456 | 62,226 | 1,013,536 | 153,746 | 92,030 | 1,062,355 | 2,383,893 | 184,946 | 3,482,562 | 197,911 |
| 1987 | 21,650 | 59,728 | 624,036 | 328,405 | 22,860 | 1,056,679 | 62,772 | 106,499 | 672,988 | 134,485 | 51,375 | 914,859 | 1,880,207 | 170,002 | 2,936,885 | 181,220 |
| 1988 | 14,740 | 108,987 | 554,480 | 180,242 | 19,053 | 877,501 | 50,244 | 37,584 | 1,040,039 | 177,532 | 119,759 | 906,598 | 2,281,512 | 185,870 | 3,159,014 | 192,541 |
| 1989 | 21,661 | 59,418 | 678,024 | 264,132 | 6,151 | 1,029,386 | 70,353 | 19,523 | 592,805 | 107,185 | 115,892 | 973,506 | 1,808,911 | 164,041 | 2,838,297 | 178,491 |
| 1990 | 21,302 | 55,516 | 555,477 | 238,818 | 14,448 | 885,561 | 56,432 | 35,159 | 439,885 | 76,456 | 96,011 | 495,498 | 1,143,009 | 93,336 | 2,028,570 | 109,069 |
| 1991 | 19,325 | 62,425 | 517,861 | 230,468 | 17,744 | 847,823 | 54,255 | 25,344 | 310,608 | 74,896 | 52,434 | 471,484 | 934,766 | 90,794 | 1,782,589 | 105,769 |
| 1992 | 31,462 | 79,113 | 396,871 | 192,028 | 19,185 | 718,658 | 43,557 | 45,949 | 496,569 | 74,984 | 107,403 | 627,853 | 1,352,759 | 119,120 | 2,071,416 | 126,833 |
| 1993 | 23,321 | 73,143 | 385,618 | 158,505 | 19,996 | 660,583 | 38,443 | 64,356 | 352,244 | 146,609 | 128,281 | 624,678 | 1,316,168 | 122,148 | 1,976,751 | 128,054 |
| 1994 | 15,698 | 50,695 | 593,827 | 174,288 | 17,134 | 851,641 | 53,830 | 50,491 | 496,896 | 204,480 | 85,743 | 660,748 | 1,498,358 | 129,716 | 2,349,998 | 140,442 |
| 1995 | 16,242 | 72,501 | 326,990 | 156,236 | 24,868 | 596,837 | 33,200 | 15,355 | 492,798 | 102,818 | 79,952 | 618,738 | 1,309,660 | 113,463 | 1,906,497 | 118,227 |
| 1996 | 26,796 | 54,588 | 224,768 | 191,239 | 15,450 | 512,841 | 29,440 | 19,103 | 431,085 | 84,459 | 82,572 | 444,657 | 1,061,876 | 87,589 | 1,574,717 | 92,404 |
| 1997 | 23,655 | 47,196 | 294,110 | 179,313 | 6,932 | 551,205 | 30,812 | 9,867 | 325,064 | 75,517 | 99,092 | 346,721 | 856,261 | 66,241 | 1,407,466 | 73,056 |
| 1998 | 28,452 | 70,799 | 369,612 | 195,926 | 4,036 | 668,824 | 33,467 | 19,163 | 372,024 | 85,090 | 213,314 | 394,098 | 1,083,688 | 77,128 | 1,752,513 | 84,076 |
| 1999 | 35,197 | 46,706 | 392,098 | 134,992 | 5,969 | 614,962 | 31,273 | 6,350 | 284,962 | 74,815 | 55,256 | 205,117 | 626,500 | 42,006 | 1,241,462 | 52,369 |
| 2000 | 36,828 | 37,300 | 600,701 | 155,941 | 11,901 | 842,672 | 51,296 | 16,861 | 361,341 | 119,103 | 80,468 | 241,550 | 819,323 | 54,887 | 1,661,995 | 75,126 |
| 10yr Av. | 25,697 | 59,447 | 410,246 | 176,894 | 14,322 | 686,605 | 39,959 | 27,284 | 392,359 | 104,277 | 98,452 | 463,564 | 1,085,936 | 90,309 | 1,772,540 | 99,636 |

Table 3.6.2.2 Estimated number of RETURNING MSW salmon by NEAC country and year

| Year | Northern Europe |  |  |  |  |  |  | Southern Europe |  |  |  |  |  |  | NEAC Area |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Finland | Iceland | Norway | Russia | Sweden | Total |  | France | Ireland | UK(EW) | UK(NI) | UK(Scot) | Total |  | Total |  |
|  |  |  |  |  |  | Est. | SD |  |  |  |  |  | Est. | SD | Est. | SD |
| 1971 | 7,296 | 37,018 | 343,756 | 204,481 | 865 | 593,416 | 53,746 | 11,255 | 174,511 | 107,660 | 15,679 | 526,864 | 835,970 | 68,716 | 1,429,385 | 87,238 |
| 1972 | 11,427 | 50,486 | 451,014 | 170,890 | 638 | 684,455 | 64,955 | 22,527 | 193,118 | 119,592 | 13,834 | 713,746 | 1,062,817 | 91,833 | 1,747,271 | 112,483 |
| 1973 | 17,421 | 52,932 | 492,614 | 285,191 | 2,043 | 850,203 | 78,742 | 13,851 | 206,349 | 110,051 | 12,032 | 789,053 | 1,131,337 | 101,633 | 1,981,540 | 128,567 |
| 1974 | 16,780 | 44,392 | 467,159 | 262,834 | 1,348 | 792,513 | 73,842 | 6,404 | 227,862 | 107,255 | 13,214 | 649,748 | 1,004,483 | 83,845 | 1,796,997 | 111,726 |
| 1975 | 16,432 | 50,875 | 440,399 | 330,259 | 326 | 838,289 | 73,364 | 12,711 | 237,614 | 121,472 | 10,822 | 717,402 | 1,100,022 | 95,753 | 1,938,311 | 120,628 |
| 1976 | 14,423 | 43,811 | 438,028 | 281,788 | 963 | 779,013 | 72,832 | 9,430 | 167,206 | 74,382 | 7,606 | 423,289 | 681,913 | 58,949 | 1,460,926 | 93,698 |
| 1977 | 13,135 | 50,659 | 425,060 | 211,489 | 745 | 701,089 | 65,681 | 7,321 | 146,713 | 81,360 | 7,293 | 513,504 | 756,191 | 67,036 | 1,457,281 | 93,850 |
| 1978 | 8,112 | 63,675 | 299,213 | 145,620 | 594 | 517,215 | 43,594 | 7,428 | 131,382 | 78,467 | 9,555 | 614,894 | 841,726 | 78,816 | 1,358,941 | 90,069 |
| 1979 | 6,459 | 41,087 | 519,467 | 160,929 | 1,680 | 729,623 | 76,810 | 8,437 | 116,739 | 59,017 | 6,775 | 506,675 | 697,643 | 65,505 | 1,427,266 | 100,949 |
| 1980 | 8,757 | 51,254 | 521,520 | 244,931 | 3,005 | 829,467 | 75,741 | 17,648 | 131,154 | 73,055 | 8,450 | 598,932 | 829,239 | 74,613 | 1,658,706 | 106,320 |
| 1981 | 13,703 | 27,068 | 547,026 | 133,546 | 865 | 722,208 | 78,673 | 12,367 | 92,108 | 86,425 | 6,673 | 634,358 | 831,930 | 79,662 | 1,554,139 | 111,962 |
| 1982 | 15,188 | 28,741 | 443,278 | 134,848 | 3,000 | 625,054 | 64,627 | 7,660 | 37,130 | 50,146 | 9,696 | 482,494 | 587,125 | 59,812 | 1,212,180 | 88,058 |
| 1983 | 16,905 | 33,537 | 429,326 | 217,029 | 2,032 | 698,828 | 48,078 | 8,147 | 97,458 | 66,396 | 13,646 | 583,971 | 769,618 | 71,687 | 1,468,446 | 86,316 |
| 1984 | 13,482 | 29,308 | 523,878 | 231,544 | 2,977 | 801,188 | 54,824 | 13,692 | 95,400 | 55,780 | 5,289 | 447,273 | 617,434 | 56,471 | 1,418,622 | 78,706 |
| 1985 | 13,775 | 22,855 | 465,102 | 257,068 | 1,171 | 759,971 | 55,410 | 10,080 | 87,521 | 57,762 | 6,885 | 499,786 | 662,034 | 64,502 | 1,422,005 | 85,034 |
| 1986 | 9,164 | 40,398 | 611,906 | 286,548 | 1,202 | 949,218 | 65,049 | 10,392 | 126,751 | 69,619 | 7,829 | 643,717 | 858,308 | 83,580 | 1,807,526 | 105,911 |
| 1987 | 12,799 | 30,982 | 441,595 | 133,956 | 3,602 | 622,934 | 42,549 | 5,506 | 104,979 | 75,617 | 4,032 | 469,756 | 659,889 | 63,594 | 1,282,823 | 76,516 |
| 1988 | 10,065 | 23,982 | 334,589 | 148,545 | 3,540 | 520,721 | 36,704 | 15,552 | 114,818 | 93,642 | 11,058 | 545,254 | 780,323 | 76,027 | 1,301,044 | 84,423 |
| 1989 | 12,421 | 21,834 | 242,252 | 180,902 | 9,644 | 467,054 | 42,590 | 7,094 | 76,214 | 68,179 | 8,906 | 475,958 | 636,351 | 64,660 | 1,103,405 | 77,427 |
| 1990 | 12,979 | 22,534 | 313,769 | 174,451 | 6,596 | 530,329 | 39,686 | 7,043 | 55,603 | 83,438 | 8,266 | 414,633 | 568,983 | 58,075 | 1,099,312 | 70,340 |
| 1991 | 14,581 | 20,209 | 265,964 | 150,390 | 7,524 | 458,668 | 32,937 | 6,399 | 39,531 | 36,014 | 4,122 | 323,069 | 409,135 | 45,312 | 867,804 | 56,018 |
| 1992 | 14,479 | 27,074 | 267,747 | 137,774 | 9,930 | 457,004 | 31,341 | 8,203 | 64,602 | 26,459 | 9,448 | 425,106 | 533,818 | 61,223 | 990,822 | 68,779 |
| 1993 | 18,787 | 22,241 | 248,269 | 204,507 | 13,187 | 506,992 | 40,237 | 3,921 | 43,427 | 41,284 | 23,036 | 383,831 | 495,500 | 55,511 | 1,002,492 | 68,560 |
| 1994 | 14,361 | 23,979 | 263,973 | 176,742 | 10,042 | 489,097 | 33,252 | 7,937 | 78,348 | 73,492 | 7,894 | 478,032 | 645,703 | 73,450 | 1,134,800 | 80,627 |
| 1995 | 12,410 | 18,615 | 279,735 | 128,620 | 6,257 | 445,638 | 30,712 | 3,799 | 78,320 | 48,095 | 6,617 | 451,708 | 588,538 | 69,586 | 1,034,176 | 76,062 |
| 1996 | 7,692 | 17,972 | 247,714 | 203,258 | 7,795 | 484,430 | 50,732 | 6,740 | 52,347 | 54,800 | 7,475 | 347,315 | 468,678 | 56,805 | 953,108 | 76,162 |
| 1997 | 12,108 | 14,576 | 169,798 | 164,279 | 5,124 | 365,885 | 40,443 | 3,480 | 34,973 | 33,681 | 9,126 | 261,367 | 342,628 | 44,625 | 708,513 | 60,225 |
| 1998 | 10,725 | 13,683 | 206,078 | 163,633 | 2,850 | 396,970 | 40,292 | 2,955 | 42,716 | 21,447 | 13,135 | 278,565 | 358,819 | 49,183 | 755,788 | 63,581 |
| 1999 | 9,645 | 22,034 | 246,263 | 162,808 | 2,392 | 443,141 | 40,226 | 6,389 | 31,823 | 40,507 | 5,850 | 223,443 | 308,012 | 40,301 | 751,153 | 56,941 |
| 2000 | 21,825 | 8,557 | 325,987 | 220,879 | 5,350 | 582,598 | 55,715 | 4,463 | 40,406 | 43,178 | 7,728 | 240,851 | 336,624 | 44,577 | 919,223 | 71,353 |
| 10yr Av. | 13,661 | 18,894 | 252,153 | 171,289 | 7,045 | 463,042 | 39,589 | 5,428 | 50,649 | 41,896 | 9,443 | 341,329 | 448,746 | 54,058 | 911,788 | 67,831 |

Table 3.6.2.3 Estimated pre-fishery abundance of MATURING 1SW salmon (potential 1SW returns) by NEAC country and year

| Year | Northern Europe |  |  |  |  |  |  | Southern Europe |  |  |  |  |  |  | NEAC Area |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Finland | Iceland | Norway | Russia | Sweden | Total |  | France | Ireland | UK(EW) | UK(NI) | UK(Scot) | Total |  | Total |  |
|  |  |  |  |  |  | Est. | SD |  |  |  |  |  | Est. | SD | Est. | SD |
| 1971 | 8,555 | 68,540 | 595,454 | 133,849 | 13,004 | 819,402 | 84,740 | 58,032 | 1,160,467 | 181,621 | 201,954 | 1,262,769 | 2,864,842 | 224,166 | 3,684,244 | 239,648 |
| 1972 | 13,209 | 69,599 | 766,284 | 146,873 | 10,493 | 1,006,458 | 108,662 | 114,289 | 1,262,444 | 201,983 | 175,472 | 1,160,483 | 2,914,670 | 212,370 | 3,921,128 | 238,555 |
| 1973 | 19,971 | 71,561 | 849,952 | 246,793 | 12,865 | 1,201,142 | 121,453 | 69,751 | 1,362,219 | 192,504 | 156,504 | 1,362,357 | 3,143,336 | 234,037 | 4,344,478 | 263,674 |
| 1974 | 18,440 | 50,556 | 798,805 | 225,642 | 18,876 | 1,112,318 | 116,893 | 32,775 | 1,495,777 | 189,376 | 169,880 | 1,423,777 | 3,311,585 | 255,914 | 4,423,903 | 281,347 |
| 1975 | 18,369 | 69,075 | 745,485 | 241,092 | 20,077 | 1,094,097 | 109,278 | 65,944 | 1,549,605 | 217,897 | 137,855 | 1,167,850 | 3,139,151 | 234,605 | 4,233,249 | 258,807 |
| 1976 | 15,800 | 57,812 | 742,753 | 181,749 | 11,400 | 1,009,514 | 107,725 | 60,388 | 1,092,174 | 136,057 | 95,151 | 1,011,217 | 2,394,986 | 197,654 | 3,404,500 | 225,104 |
| 1977 | 14,256 | 69,105 | 728,229 | 153,289 | 5,367 | 970,246 | 106,015 | 46,734 | 964,895 | 155,424 | 94,063 | 1,093,431 | 2,354,547 | 201,232 | 3,324,793 | 227,450 |
| 1978 | 10,401 | 84,138 | 786,092 | 166,448 | 6,303 | 1,053,382 | 112,601 | 47,487 | 862,292 | 153,254 | 121,473 | 1,191,559 | 2,376,064 | 217,114 | 3,429,446 | 244,576 |
| 1979 | 10,793 | 74,719 | 890,685 | 190,866 | 6,518 | 1,173,581 | 126,939 | 53,982 | 768,743 | 119,268 | 86,631 | 1,023,396 | 2,052,020 | 174,127 | 3,225,601 | 215,485 |
| 1980 | 10,785 | 38,216 | 894,905 | 126,126 | 8,238 | 1,078,269 | 126,081 | 113,101 | 589,195 | 151,787 | 111,163 | 638,362 | 1,603,610 | 118,156 | 2,681,879 | 172,793 |
| 1981 | 10,143 | 53,745 | 619,880 | 91,294 | 14,755 | 789,817 | 86,854 | 90,260 | 405,400 | 185,179 | 87,741 | 799,075 | 1,567,656 | 128,929 | 2,357,473 | 155,455 |
| 1982 | 7,337 | 35,477 | 452,987 | 138,948 | 13,198 | 647,947 | 65,063 | 55,626 | 726,841 | 110,569 | 125,702 | 1,113,661 | 2,132,400 | 185,169 | 2,780,346 | 196,267 |
| 1983 | 10,973 | 56,086 | 749,857 | 201,532 | 17,304 | 1,035,751 | 68,539 | 60,263 | 1,193,027 | 150,140 | 175,983 | 1,195,733 | 2,775,147 | 212,602 | 3,810,898 | 223,377 |
| 1984 | 13,441 | 35,569 | 945,864 | 191,098 | 24,367 | 1,210,339 | 84,688 | 98,591 | 569,280 | 127,880 | 69,192 | 1,222,602 | 2,087,546 | 199,781 | 3,297,885 | 216,990 |
| 1985 | 16,850 | 62,259 | 993,557 | 251,776 | 29,023 | 1,353,466 | 92,460 | 36,429 | 1,141,531 | 134,982 | 89,538 | 921,675 | 2,324,155 | 174,199 | 3,677,621 | 197,216 |
| 1986 | 16,849 | 90,393 | 832,162 | 220,832 | 31,367 | 1,191,602 | 76,786 | 67,656 | 1,098,947 | 167,027 | 99,879 | 1,148,032 | 2,581,542 | 202,850 | 3,773,144 | 216,897 |
| 1987 | 23,622 | 64,725 | 676,312 | 356,207 | 24,954 | 1,145,819 | 68,117 | 115,541 | 729,702 | 146,059 | 55,830 | 988,938 | 2,036,070 | 185,476 | 3,181,889 | 197,588 |
| 1988 | 16,149 | 118,113 | 601,178 | 195,683 | 20,845 | 951,968 | 54,601 | 40,900 | 1,127,516 | 192,738 | 129,862 | 980,405 | 2,471,422 | 203,319 | 3,423,390 | 210,523 |
| 1989 | 23,587 | 64,397 | 734,752 | 286,500 | 6,788 | 1,116,024 | 77,167 | 21,262 | 642,728 | 116,400 | 125,627 | 1,052,198 | 1,958,215 | 178,566 | 3,074,239 | 194,527 |
| 1990 | 23,170 | 60,170 | 601,683 | 259,016 | 15,753 | 959,791 | 61,124 | 38,172 | 476,938 | 83,041 | 104,026 | 535,772 | 1,237,949 | 101,933 | 2,197,740 | 118,855 |
| 1991 | 20,993 | 67,660 | 560,807 | 249,868 | 19,302 | 918,630 | 59,123 | 27,520 | 336,708 | 81,268 | 56,820 | 509,870 | 1,012,186 | 98,803 | 1,930,816 | 115,142 |
| 1992 | 34,116 | 85,763 | 429,806 | 208,204 | 20,833 | 778,722 | 47,946 | 49,800 | 538,275 | 81,333 | 116,269 | 678,610 | 1,464,287 | 130,142 | 2,243,009 | 138,693 |
| 1993 | 25,296 | 79,278 | 417,606 | 171,833 | 21,703 | 715,717 | 42,349 | 69,809 | 381,796 | 158,921 | 138,864 | 675,103 | 1,424,494 | 132,927 | 2,140,211 | 139,510 |
| 1994 | 17,033 | 54,948 | 643,007 | 188,947 | 18,591 | 922,526 | 59,475 | 54,734 | 538,583 | 221,668 | 92,830 | 713,908 | 1,621,723 | 141,363 | 2,544,248 | 153,365 |
| 1995 | 17,628 | 78,573 | 354,108 | 169,380 | 26,978 | 646,668 | 36,239 | 16,666 | 534,157 | 111,481 | 86,562 | 668,484 | 1,417,349 | 123,241 | 2,064,017 | 128,459 |
| 1996 | 29,057 | 59,170 | 243,609 | 207,326 | 16,779 | 555,940 | 33,029 | 20,719 | 467,261 | 91,596 | 89,415 | 480,362 | 1,149,354 | 95,770 | 1,705,293 | 101,305 |
| 1997 | 25,632 | 51,160 | 318,409 | 194,373 | 7,519 | 597,093 | 34,240 | 10,692 | 352,341 | 81,868 | 107,241 | 374,631 | 926,772 | 72,714 | 1,523,866 | 80,372 |
| 1998 | 30,833 | 76,730 | 400,155 | 212,347 | 4,380 | 724,445 | 37,035 | 20,761 | 403,206 | 92,224 | 230,849 | 425,708 | 1,172,748 | 84,270 | 1,897,193 | 92,049 |
| 1999 | 38,140 | 50,618 | 424,487 | 146,306 | 6,473 | 666,023 | 34,749 | 6,880 | 308,851 | 81,079 | 59,806 | 221,592 | 678,208 | 46,272 | 1,344,231 | 57,867 |
| 2000 | 39,914 | 40,425 | 650,043 | 169,009 | 12,906 | 912,297 | 55,036 | 18,278 | 391,618 | 129,087 | 87,080 | 260,984 | 887,048 | 60,316 | 1,799,345 | 81,651 |
| 10yr Av. | 27,864 | 64,432 | 444,204 | 191,759 | 15,546 | 743,806 | 43,922 | 29,586 | 425,280 | 113,052 | 106,574 | 500,925 | 1,175,417 | 98,582 | 1,919,223 | 108,841 |

Table 3.6.2.4 Estimated pre-fishery abundance of NON-MATURING 1SW salmon (potential MSW returns) by NEAC country and year

| Year | Northern Europe |  |  |  |  |  |  | Southern Europe |  |  |  |  |  |  | NEAC Area <br> Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Finland | Iceland | Norway | Russia | Sweden | Total |  | France | Ireland | UK(EW) | UK(NI) | UK(Scot) | Total |  |  |  |
|  |  |  |  |  |  | Est. | SD |  |  |  |  |  | Est. | SD | Est. | SD |
| 1971 | 14,088 | 68,540 | 548,639 | 123,242 | 11,868 | 766,377 | 79,972 | 40,798 | 312,903 | 223,928 | 16,394 | 1,054,004 | 1,648,028 | 101,101 | 2,414,405 | 128,907 |
| 1972 | 24,988 | 69,599 | 7J6,255 | 135,241 | 9,546 | 945,629 | 87,414 | 26,541 | 307,809 | 190,913 | 14,246 | 1,100,654 | 1,640,164 | 112,952 | 2,585,793 | 142,827 |
| 1973 | 23,274 | 71,561 | 733,174 | 227,366 | 11,712 | 1,117,087 | 83,846 | 16,804 | 326,456 | 181,741 | 15,641 | 913,330 | 1,453,970 | 90,464 | 2,571,056 | 123,345 |
| 1974 | 23,313 | 50,556 | 736,277 | 207,967 | 17,307 | 1,035,419 | 79,134 | 23,891 | 337,147 | 197,185 | 12,835 | 996,096 | 1,567,154 | 110,307 | 2,602,573 | 135,756 |
| 1975 | 20,274 | 69,075 | 636,855 | 222,157 | 18,386 | 1,016,747 | 78,263 | 21,874 | 261,486 | 150,052 | 9,008 | 656,130 | 1,098,549 | 61,180 | 2,115,296 | 99,338 |
| 1976 | 17,678 | 57,812 | 634,807 | 167,519 | 10,434 | 938,250 | 76,317 | 15,037 | 212,149 | 133,975 | 8,641 | 705,172 | 1,074,974 | 73,343 | 2,013,224 | 105,847 |
| 1977 | 11,467 | 69,105 | 671,580 | 141,326 | 4,888 | 898,364 | 50,840 | 16,086 | 198,131 | 135,267 | 11,317 | 833,365 | 1,194,166 | 90,700 | 2,092,530 | 103,977 |
| 1978 | 9,904 | 84,138 | 724,892 | 153,472 | 5,766 | 978,172 | 91,331 | 14,650 | 168,043 | 98,039 | 8,022 | 681,071 | 969,825 | 76,088 | 1,947,996 | 118,873 |
| 1979 | 15,576 | 74,719 | 821,339 | 175,948 | 5,927 | 1,093,508 | 87,494 | 27,811 | 203,931 | 129,841 | 10,008 | 850,845 | 1,222,435 | 92,224 | 2,315,944 | 127,124 |
| 1980 | 24,376 | 38,216 | 824,362 | 115,940 | 7,378 | 1,010,272 | 96,396 | 21,029 | 162,225 | 146,040 | 7,904 | 923,147 | 1,260,345 | 95,330 | 2,270,617 | 135,573 |
| 1981 | 25,588 | 53,745 | 539,815 | 83,525 | 13,250 | 745,923 | 77,631 | 14,463 | 90,592 | 96,785 | 11,483 | 724,774 | 938,097 | 70,458 | 1,684,020 | 104,837 |
| 1982 | 26,048 | 35,477 | 415,943 | 127,541 | 11,852 | 616,860 | 46,162 | 13,998 | 152,954 | 109,055 | 16,169 | 815,518 | 1,107,694 | 81,858 | 1,724,555 | 93,977 |
| 1983 | 20,032 | 56,086 | 630,376 | 185,222 | 15,598 | 967,315 | 55,115 | 18,222 | 133,454 | 81,608 | 6,264 | 603,617 | 843,166 | 62,287 | 1,810,480 | 83,171 |
| 1984 | 20,401 | 35,569 | 872,447 | 175,950 | 22,290 | 1,126,657 | 47,336 | 13,540 | 122,029 | 81,789 | 8,155 | 662,061 | 887,573 | 68,353 | 2,014,230 | 83,143 |
| 1985 | 15,568 | 62,259 | 916,774 | 231,986 | 26,626 | 1,253,213 | 61,571 | 17,286 | 187,666 | 114,919 | 9,273 | 877,484 | 1,206,627 | 93,626 | 2,459,839 | 112,057 |
| 1986 | 19,731 | 90,393 | 737,763 | 203,397 | 28,737 | 1,110,021 | 46,934 | 11,002 | 159,016 | 119,408 | 4,777 | 662,991 | 957,193 | 68,051 | 2,067,214 | 82,666 |
| 1987 | 14,655 | 64,725 | 624,036 | 328,405 | 22,860 | 1,054,680 | 36,761 | 22,576 | 164,306 | 137,277 | 13,095 | 727,119 | 1,064,373 | 75,695 | 2,119,053 | 84,150 |
| 1988 | 18,242 | 118,113 | 554,480 | 180,242 | 19,053 | 890,129 | 26,444 | 13,679 | 126,311 | 113,699 | 10,550 | 665,920 | 930,157 | 66,930 | 1,820,286 | 71,964 |
| 1989 | 18,468 | 64,397 | 678,024 | 264,132 | 6,151 | 1,031,172 | 34,490 | 10,056 | 82,370 | 112,302 | 9,786 | 549,856 | 764,372 | 61,055 | 1,795,544 | 70,123 |
| 1990 | 18,618 | 60,170 | 555,477 | 238,818 | 14,448 | 887,532 | 28,867 | 8,431 | 54,496 | 49,021 | 4,883 | 410,251 | 527,083 | 47,080 | 1,414,614 | 55,226 |
| 1991 | 17,962 | 67,660 | 517,861 | 230,468 | 17,744 | 851,694 | 27,964 | 11,795 | 89,195 | 43,581 | 11,189 | 538,287 | 694,046 | 64,463 | 1,545,741 | 70,267 |
| 1992 | 22,911 | 85,763 | 336,871 | 192,028 | 19,185 | 716,757 | 27,116 | 5,930 | 59,707 | 57,055 | 27,286 | 479,045 | 629,023 | 59,183 | 1,345,780 | 65,099 |
| 1993 | 17,596 | 79,278 | 335,618 | 158,505 | 19,996 | 660,994 | 22,234 | 9,441 | 94,509 | 88,852 | 9,352 | 577,555 | 779,709 | 79,297 | 1,440,702 | 82,355 |
| 1994 | 15,358 | 54,948 | 593,827 | 174,288 | 17,134 | 855,554 | 27,532 | 4,529 | 94,634 | 58,377 | 7,834 | 547,025 | 712,399 | 79,293 | 1,567,953 | 83,937 |
| 1995 | 9,749 | 78,573 | 326,990 | 156,236 | 24,868 | 596,416 | 25,437 | 8,346 | 65,442 | 68,209 | 8,856 | 425,597 | 576,449 | 62,187 | 1,172,865 | 67,188 |
| 1996 | 14,447 | 59,170 | 224,768 | 191,239 | 15,450 | 505,075 | 17,509 | 4,433 | 43,256 | 42,089 | 10,806 | 315,978 | 416,561 | 48,592 | 921,636 | 51,650 |
| 1997 | 12,786 | 51,160 | 294,110 | 179,313 | 6,932 | 544,300 | 20,964 | 3,658 | 51,597 | 26,572 | 15,562 | 334,549 | 431,938 | 51,043 | 976,239 | 55,181 |
| 1998 | 11,463 | 76,730 | 339,612 | 195,926 | 4,036 | 657,767 | 25,068 | 7,611 | 37,960 | 48,756 | 6,926 | 267,167 | 368,421 | 44,133 | 1,026,187 | 50,756 |
| 1999 | 25,943 | 50,618 | 332,098 | 134,992 | 5,969 | 609,619 | 30,379 | 5,333 | 48,209 | 52,035 | 9,155 | 288,260 | 403,052 | 45,888 | 1,012,671 | 55,033 |
| 10yr Av. | 16,683 | 66,407 | 475,723 | 185,181 | 14,576 | 688,571 | 25,307 | 6,951 | 63,906 | 53,455 | 11,185 | 418,371 | 553,868 | 58,116 | 1,242,439 | 63,669 |

Table 3.6.2.5 Estimated number of 1 SW SPAWNERS by NEAC country and year

| Year | Northern Europe |  |  |  |  |  |  | Southern Europe |  |  |  |  |  |  | NEAC Area |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Finland | Iceland | Norway | Russia | Sweden | Total |  | France | Ireland | UK(EW) | UK(N) | UK(Scot) | Total |  |  |  |
|  |  |  |  |  |  | Est. | SD |  |  |  |  |  | Est. | SD | Est. | SD |
| eggs/F | 5,000 | 4,500 | 4,500 | 4,500 | 3,000 | 4403 |  | 3,450 | 3,400 | 3,000 | 3,400 | 5,000 | 4490 |  | 2623 |  |
| \% Fem | 12\% | 45\% | 45\% | 45\% | 50\% | 47\% |  | 77\% | 60\% | 45\% | 60\% | 40\% | 45\% |  | 27\% |  |
| 1971 | 3,618 | 32,155 | 112,989 | 62,664 | 2,159 | 181,431 | 57,145 | 51,687 | 302,707 | 114,127 | 38,082 | 807,337 | 1,313,940 | 194,913 | 1,527,527 | 203,118 |
| 1972 | 5,557 | 32,456 | 143,226 | 68,357 | 1,688 | 218,827 | 74,199 | 101,879 | 325,211 | 126,427 | 32,455 | 730,071 | 1,316,042 | 182,173 | 1,567,325 | 196,703 |
| 1973 | 8,399 | 33,444 | 159,750 | 115,174 | 2,059 | 285,382 | 82,176 | 62,081 | 351,827 | 120,585 | 29,833 | 859,635 | 1,423,961 | 201,376 | 1,742,788 | 217,498 |
| 1974 | 7,821 | 23,570 | 149,441 | 105,098 | 3,170 | 265,529 | 80,102 | 29,133 | 385,428 | 118,519 | 31,248 | 920,399 | 1,484,727 | 220,848 | 1,773,826 | 234,925 |
| 1975 | 7,825 | 32,227 | 139,390 | 112,326 | 3,394 | 262,935 | 74,029 | 58,760 | 399,322 | 136,379 | 25,682 | 773,116 | 1,393,259 | 198,582 | 1,688,420 | 211,932 |
| 1976 | 6,722 | 26,890 | 138,127 | 84,457 | 1,817 | 231,123 | 74,024 | 53,803 | 280,702 | 84,108 | 17,515 | 674,066 | 1,110,194 | 171,875 | 1,368,208 | 187,138 |
| 1977 | 6,096 | 32,466 | 138,958 | 71,963 | 949 | 217,967 | 72,729 | 41,666 | 250,728 | 95,835 | 17,817 | 742,353 | 1,148,400 | 177,887 | 1,398,833 | 192,180 |
| 1978 | 4,429 | 39,193 | 146,808 | 77,491 | 1,065 | 229,795 | 77,269 | 42,350 | 221,977 | 92,900 | 22,361 | 818,144 | 1,197,733 | 193,991 | 1,466,721 | 208,814 |
| 1979 | 4,553 | 34,956 | 167,840 | 89,202 | 1,065 | 262,660 | 85,832 | 48,079 | 198,718 | 71,700 | 16,087 | 690,183 | 1,024,767 | 153,981 | 1,322,383 | 176,287 |
| 1980 | 4,501 | 17,879 | 168,590 | 58,791 | 1,294 | 233,176 | 85,198 | 100,717 | 144,168 | 86,829 | 20,707 | 439,025 | 791,446 | 103,586 | 1,042,501 | 134,122 |
| 1981 | 4,173 | 25,322 | 118,454 | 42,648 | 2,318 | 167,594 | 58,119 | 80,239 | 93,324 | 105,282 | 16,279 | 551,670 | 846,794 | 116,182 | 1,039,710 | 129,908 |
| 1982 | 2,979 | 16,654 | 85,668 | 77,861 | 2,150 | 168,657 | 44,755 | 49,326 | 171,802 | 61,874 | 23,589 | 771,854 | 1,078,445 | 166,208 | 1,263,756 | 172,128 |
| 1983 | 4,539 | 26,259 | 162,735 | 112,832 | 2,826 | 282,931 | 48,218 | 53,463 | 293,850 | 83,093 | 32,493 | 843,788 | 1,306,686 | 184,626 | 1,615,877 | 190,818 |
| 1984 | 5,685 | 16,617 | 200,604 | 106,996 | 4,035 | 317,320 | 60,257 | 87,842 | 133,372 | 69,424 | 12,769 | 864,860 | 1,168,267 | 179,900 | 1,502,203 | 189,723 |
| 1985 | 7,137 | 29,040 | 217,418 | 140,906 | 4,912 | 370,373 | 67,222 | 32,378 | 252,505 | 72,085 | 16,445 | 656,429 | 1,029,841 | 147,641 | 1,429,254 | 162,224 |
| 1986 | 7,175 | 42,066 | 181,692 | 123,248 | 5,322 | 317,437 | 56,299 | 58,826 | 210,026 | 88,803 | 18,203 | 811,736 | 1,187,593 | 175,625 | 1,547,097 | 184,428 |
| 1987 | 10,019 | 30,199 | 153,372 | 206,629 | 4,318 | 374,339 | 53,869 | 100,499 | 156,203 | 78,058 | 16,189 | 709,864 | 1,060,813 | 164,133 | 1,465,350 | 172,747 |
| 1988 | 6,820 | 55,225 | 136,877 | 113,569 | 3,400 | 260,667 | 43,060 | 35,484 | 231,810 | 103,905 | 42,575 | 720,005 | 1,133,781 | 173,429 | 1,449,673 | 178,695 |
| 1989 | 7,787 | 30,138 | 232,306 | 166,498 | 1,143 | 407,735 | 63,774 | 18,423 | 153,411 | 60,362 | 13,500 | 774,626 | 1,020,341 | 159,641 | 1,458,214 | 171,908 |
| 1990 | 7,804 | 28,175 | 196,060 | 150,593 | 2,676 | 357,134 | 50,560 | 33,259 | 147,612 | 42,586 | 37,160 | 402,882 | 663,500 | 90,258 | 1,048,809 | 103,454 |
| 1991 | 7,010 | 31,783 | 177,050 | 162,266 | 3,317 | 349,644 | 48,184 | 23,944 | 107,278 | 42,536 | 18,913 | 386,876 | 579,547 | 89,118 | 960,974 | 101,310 |
| 1992 | 11,433 | 40,257 | 132,172 | 135,177 | 3,577 | 282,359 | 39,692 | 43,449 | 180,065 | 42,990 | 47,904 | 511,838 | 826,246 | 116,188 | 1,148,863 | 122,781 |
| 1993 | 8,430 | 37,256 | 131,235 | 111,613 | 3,277 | 254,555 | 35,803 | 60,756 | 100,868 | 90,941 | 76,071 | 516,622 | 845,258 | 120,252 | 1,137,069 | 125,469 |
| 1994 | 5,699 | 25,817 | 209,147 | 122,714 | 4,769 | 342,329 | 48,689 | 47,691 | 137,626 | 128,803 | 26,177 | 547,649 | 887,946 | 126,123 | 1,256,091 | 135,194 |
| 1995 | 5,852 | 36,774 | 112,737 | 109,835 | 9,711 | 238,136 | 30,700 | 13,686 | 128,585 | 66,256 | 26,713 | 516,477 | 751,717 | 109,746 | 1,026,627 | 113,959 |
| 1996 | 12,435 | 27,777 | 71,852 | 134,604 | 6,036 | 224,929 | 27,615 | 17,040 | 121,123 | 57,641 | 35,757 | 369,417 | 600,977 | 84,197 | 853,683 | 88,610 |
| 1997 | 10,997 | 23,899 | 110,598 | 125,987 | 2,589 | 250,171 | 29,360 | 8,807 | 96,019 | 52,301 | 40,592 | 293,228 | 490,947 | 65,742 | 765,018 | 72,000 |
| 1998 | 13,201 | 35,743 | 134,949 | 137,515 | 1,164 | 286,830 | 31,163 | 17,098 | 112,674 | 60,350 | 160,340 | 333,459 | 683,921 | 76,595 | 1,006,494 | 82,692 |
| 1999 | 12,688 | 23,724 | 145,123 | 94,955 | 1,763 | 254,530 | 28,289 | 5,660 | 88,681 | 55,532 | 20,481 | 176,536 | 346,890 | 41,474 | 625,144 | 50,204 |
| 2000 | 13,463 | 18,960 | 220,993 | 109,725 | 3,564 | 347,745 | 46,961 | 15,069 | 113,004 | 89,767 | 33,905 | 210,858 | 462,603 | 54,216 | 829,309 | 71,727 |
| 10yr.av. | 9,065 | 31,762 | 150,811 | 135,680 | 3,826 | 299,382 |  | 28,415 | 128,526 | 64,479 | 48,313 | 465,307 | 735,040 |  | 1,066,184 |  |

Table 3.6.2.6 Estimated number of MSW SPAWNERS by NEAC country and year

| Year | Northern Europe |  |  |  |  |  |  | Southern Europe |  |  |  |  |  |  | NEAC Area |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Finland | Iceland | Norway | Russia | Sweden | Total |  | France | Ireland | UK(EW) | UK(NI) | UK(Scot) | Total |  | Total |  |
|  |  |  |  |  |  | Est. | SD |  |  |  |  |  | Est. | SD |  | SD |
| eggs/F | 13,000 | 10,500 | 10,500 | 10,500 | 6,000 |  |  | 6,900 | 7,000 | 6,000 | 6,000 | 10,000 |  |  |  |  |
| \% Fem | 77\% | 80\% | 80\% | 80\% | 70\% |  |  | 77\% | 85\% | 70\% | 85\% | 60\% |  |  |  |  |
| 1971 | 3,083 | 14,956 | 69,680 | 103,303 | 211 | 176,276 | 43,246 | 7,195 | 85,589 | 79,157 | 7,842 | 304,255 | 484,038 | 66,554 | 675,270 | 79,370 |
| 1972 | 4,835 | 20,436 | 91,802 | 86,456 | 174 | 183,268 | 50,153 | 14,407 | 95,038 | 87,985 | 6,966 | 413,329 | 617,724 | 89,149 | 821,428 | 102,288 |
| 1973 | 7,422 | 21,392 | 100,055 | 144,145 | 462 | 252,084 | 64,271 | 8,881 | 101,497 | 80,978 | 6,035 | 462,306 | 659,697 | 98,689 | 933,172 | 117,772 |
| 1974 | 7,100 | 18,024 | 95,746 | 133,315 | 341 | 236,502 | 59,258 | 4,094 | 112,261 | 79,069 | 6,619 | 390,423 | 592,467 | 81,268 | 846,993 | 100,578 |
| 1975 | 6,875 | 20,663 | 90,273 | 167,537 | 79 | 264,764 | 60,469 | 8,091 | 117,075 | 89,530 | 5,440 | 432,568 | 652,704 | 93,069 | 938,131 | 110,988 |
| 1976 | 6,149 | 17,819 | 89,963 | 143,134 | 221 | 239,468 | 58,900 | 6,050 | 82,403 | 54,439 | 3,825 | 264,760 | 411,476 | 57,414 | 668,763 | 82,253 |
| 1977 | 5,672 | 20,599 | 87,322 | 107,395 | 189 | 200,578 | 51,771 | 4,721 | 72,311 | 58,997 | 3,634 | 322,646 | 462,309 | 65,367 | 683,486 | 83,385 |
| 1978 | 3,455 | 25,733 | 60,530 | 73,555 | 159 | 137,700 | 33,326 | 4,763 | 64,593 | 56,167 | 4,772 | 391,081 | 521,376 | 77,123 | 684,809 | 84,016 |
| 1979 | 3,396 | 16,549 | 102,838 | 81,104 | 417 | 189,756 | 58,787 | 5,382 | 57,364 | 41,834 | 3,411 | 324,813 | 432,805 | 64,082 | 639,110 | 86,962 |
| 1980 | 4,617 | 20,717 | 105,602 | 123,751 | 789 | 234,758 | 58,340 | 11,278 | 64,488 | 49,860 | 4,226 | 386,333 | 516,185 | 73,657 | 771,660 | 93,962 |
| 1981 | 7,191 | 10,994 | 112,054 | 67,750 | 216 | 187,210 | 58,913 | 8,287 | 37,881 | 58,672 | 3,354 | 418,227 | 526,420 | 78,778 | 724,625 | 98,371 |
| 1982 | 7,983 | 11,566 | 88,985 | 81,640 | 752 | 179,359 | 49,433 | 5,140 | 16,295 | 33,573 | 4,848 | 323,695 | 383,551 | 59,328 | 574,476 | 77,223 |
| 1983 | 8,837 | 13,536 | 96,805 | 131,593 | 487 | 237,722 | 40,613 | 5,447 | 48,243 | 44,126 | 6,865 | 402,830 | 507,511 | 71,014 | 758,769 | 81,807 |
| 1984 | 7,177 | 11,928 | 122,965 | 141,157 | 802 | 272,101 | 47,439 | 9,252 | 62,092 | 36,658 | 2,635 | 314,096 | 424,734 | 56,041 | 708,762 | 73,423 |
| 1985 | 7,219 | 9,286 | 107,235 | 156,615 | 271 | 271,341 | 49,028 | 6,750 | 54,563 | 37,468 | 3,438 | 357,543 | 459,761 | 64,091 | 740,388 | 80,693 |
| 1986 | 4,840 | 16,419 | 151,054 | 174,572 | 302 | 330,767 | 56,783 | 6,992 | 79,083 | 45,091 | 3,935 | 460,343 | 595,443 | 83,001 | 942,629 | 100,566 |
| 1987 | 6,766 | 12,608 | 109,054 | 81,709 | 894 | 198,423 | 35,835 | 3,706 | 63,679 | 49,139 | 2,204 | 340,943 | 459,671 | 63,137 | 670,701 | 72,598 |
| 1988 | 5,391 | 9,640 | 80,955 | 89,868 | 912 | 177,126 | 32,260 | 10,552 | 68,853 | 60,666 | 7,072 | 405,363 | 552,506 | 75,539 | 739,271 | 82,139 |
| 1989 | 5,303 | 8,823 | 82,075 | 141,562 | 2,344 | 231,285 | 41,131 | 4,794 | 38,875 | 42,945 | 3,578 | 357,786 | 447,979 | 64,409 | 688,087 | 76,422 |
| 1990 | 5,506 | 9,110 | 107, 176 | 140,193 | 1,655 | 254,530 | 36,914 | 4,743 | 30,729 | 52,075 | 5,161 | 314,981 | 407,689 | 57,890 | 671,330 | 68,657 |
| 1991 | 6,155 | 8,209 | 91,101 | 120,994 | 1,824 | 220,074 | 30,529 | 4,299 | 22,321 | 22,836 | 2,362 | 248,852 | 300,669 | 45,213 | 528,952 | 54,555 |
| 1992 | 6,092 | 10,980 | 90,336 | 110,798 | 2,467 | 209,692 | 29,151 | 5,503 | 37,713 | 16,894 | 6,327 | 330,991 | 397,429 | 61,095 | 618,102 | 67,693 |
| 1993 | 7,845 | 9,012 | 82,114 | 164,447 | 3,179 | 257,585 | 38,747 | 2,621 | 20,414 | 28,087 | 20,286 | 302,609 | 374,016 | 55,368 | 640,613 | 67,579 |
| 1994 | 6,038 | 9,786 | 88,637 | 142,345 | 2,662 | 239,682 | 31,775 | 5,637 | 45,581 | 50,794 | 4,746 | 381,179 | 487,937 | 73,222 | 737,405 | 79,819 |
| 1995 | 5,258 | 7,558 | 95,188 | 103,469 | 1,880 | 205,795 | 28,630 | 2,704 | 45,258 | 33,887 | 3,816 | 363,192 | 448,857 | 69,445 | 662,211 | 75,115 |
| 1996 | 4,065 | 7,288 | 82,037 | 173,972 | 2,241 | 264,314 | 49,947 | 4,798 | 28,023 | 40,370 | 5,015 | 281,922 | 360,129 | 56,757 | 631,731 | 75,605 |
| 1997 | 6,413 | 5,884 | 61,359 | 140,456 | 1,574 | 209,803 | 39,926 | 2,479 | 16,374 | 25,120 | 6,062 | 215,611 | 265,646 | 44,601 | 481,333 | 59,861 |
| 1998 | 5,710 | 5,554 | 76,111 | 140,054 | 833 | 222,708 | 39,595 | 2,109 | 21,658 | 16,396 | 10,350 | 233,888 | 284,401 | 49,163 | 512,663 | 63,125 |
| 1999 | 4,506 | 8,938 | 88,503 | 139,284 | 676 | 232,969 | 39,106 | 4,558 | 15,822 | 32,233 | 4,011 | 188,843 | 245,467 | 40,283 | 487,374 | 56,143 |
| 2000 | 10,202 | 3,471 | 117,750 | 189,057 | 1,544 | 318,552 | 54,619 | 3,186 | 20,131 | 34,390 | 5,267 | 207,119 | 270,092 | 44,562 | 592,115 | 70,491 |
| 10yr.av. | 6,228 | 7,668 | 87,514 | 142,488 | 1,888 | 238,117 |  | 3,789 | 27,329 | 30,101 | 6,824 | 275,421 | 343,464 |  | 589,250 |  |

Table 3.6.6.1 Effects of taking account of fish farm escapees in Norwegian PFA assessment.

|  | Numbers estimated from PFA and CL analysis |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Maturing 1SW recruits | Non-mat. <br> 1SW recruits | Total 1SW recruits | 1SW spawners | MSW spawners | National CL |
| NOT taking account of farmed fish | 510,398 | 341,371 | 851,770 | 185,741 | 104,690 | 191,000 |
| Taking account of farm fish | 464,109 | 290,096 | 754,206 | 182,359 | 100,588 | 183,000 |
| Difference | 46,289 | 51,275 | 97,564 | 3,382 | 4,103 | 8,000 |
| \% difference | 10\% | 18\% | 13\% | 2\% | 4\% | 4\% |

Table 3.6.7.1 Sensitivity of Pre-Fishery Abundance estimates for 1SW and MSW stocks in Northern and Southern Europe to changes in input data to run-reconstruction model. [Based upon input data used in 2001 assessment]

| Country | Sea age | Input data for 2001 assessment |  |  |  |  | Recruits (PFA) | Effect of changing:- |  |  | Effect of changing: |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Catch | Non-rep' rate | Exploit'n rate | Extra catch | Time |  | Non-rep' rate by by adding $\qquad$ 0.1 | Exp'n rate by by adding 0.1 | $\begin{gathered} \text { Time } \\ \text { by } \\ \text { by adding } \\ 2.0 \\ \hline \end{gathered}$ | Non-rep' rate by multiplying 1.2 | Exploit'n rate by multiplying 1.2 | Time by multiplying 1.2 |
| Northern European Stock Complex - 1SW |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Iceland 1 | 1SW | 11,894 | 0.02 | 0.50 |  | 8.0 | 26,295 | 0.3\% | -0.5\% | 0.1\% | 0.0\% | -0.5\% | 0.0\% |
| Iceland 2 | 15W | 6,081 | 0.02 | 0.50 |  | 8.0 | 13,444 | 0.2\% | -0.2\% | 0.0\% | 0.0\% | -0.2\% | 0.0\% |
| Finland | 1SW | 17,499 | 0.25 | 0.65 |  | 8.0 | 38,885 | 0.7\% | -0.6\% | 0.1\% | 0.3\% | -0.7\% | 0.1\% |
| Norway-N | 15W | 77,121 | 0.35 | 0.70 |  | 8.0 | 183,614 | 3.6\% | -2.5\% | 0.4\% | 2.4\% | -3.3\% | 0.3\% |
| Norway-M | 15W | 85,179 | 0.35 | 0.60 |  | 8.0 | 236,598 | 4.7\% | -3.7\% | 0.5\% | 3.1\% | -4.3\% | 0.4\% |
| Norway-S | 15W | 85,179 | 0.35 | 0.60 |  | 8.0 | 236,598 | 4.7\% | -3.7\% | 0.5\% | 3.1\% | -4.3\% | 0.4\% |
| Sweden | 15W | 7,103 | 0.15 | 0.73 |  | 8.0 | 12,486 | 0.2\% | -0.2\% | 0.0\% | 0.0\% | -0.2\% | 0.0\% |
| Russia | 1SW | 27,702 | 0.40 | 0.30 |  | 8.0 | 166,718 | 3.6\% | -4.6\% | 0.4\% | 2.8\% | -3.0\% | 0.3\% |
| Faroes | 15W | 225 | 0.15 | 1.00 |  | 0.5 | 266 | - | - | - | - | - | - |
| Total Northern Area - 1SW: |  |  |  |  |  |  | 914,904 |  |  |  |  |  |  |
| Northern European Stock Complex -MSW |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Iceland 1 | MSW | 2,358 | 0.02 | 0.60 |  | 17.0 | 4,753 | 0.1\% | -0.1\% | 0.0\% | 0.0\% | -0.1\% | 0.0\% |
| Iceland 2 | MSW | 2,624 | 0.02 | 0.60 |  | 17.0 | 5,290 | 0.1\% | -0.1\% | 0.0\% | 0.0\% | -0.1\% | 0.0\% |
| Finland | MSW | 8,698 | 0.25 | 0.55 |  | 17.0 | 24,993 | 0.6\% | -0.6\% | 0.1\% | 0.3\% | -0.6\% | 0.1\% |
| Norway-N | MSW | 57,286 | 0.35 | 0.70 |  | 17.0 | 149,234 | 4.0\% | -2.8\% | 0.4\% | 2.7\% | -3.7\% | 0.8\% |
| Norway-M | MSW | 38,204 | 0.35 | 0.60 |  | 17.0 | 116,111 | 3.1\% | -2.5\% | 0.3\% | 2.1\% | -2.9\% | 0.6\% |
| Norway-S | MSW | 38,204 | 0.35 | 0.60 |  | 17.0 | 116,111 | 3.1\% | -2.5\% | 0.3\% | 2.1\% | -2.9\% | 0.6\% |
| Sweden | MSW | 3,196 | 0.15 | 0.73 |  | 17.0 | 6,147 | 0.1\% | -0.1\% | 0.0\% | 0.0\% | -0.2\% | 0.0\% |
| Russia | MSW | 9.413 | 0.70 | 0.15 |  | 17.0 | 247,939 | 18.4\% | -14.8\% | 0.7\% | 32.3\% | 6.1\% | 1.3\% |
| Faroes | MSW | 1,765 | 0.00 | 1.00 |  | 1.5 | 1,792 | - | - | - | - | - | - |
| Total Northern Area - MSW: |  |  |  |  |  |  | 672,371 |  |  |  |  |  |  |
| Southern European Stock Complex - 1SW |  |  |  |  |  |  |  |  |  |  |  |  |  |
| France | 15W | 1,792 | 0.00 | 0.13 |  | 8.0 | 15,530 | 0.2\% | -0.8\% | 0.0\% | 0.0\% | -0.3\% | 0.0\% |
| Ireland | 15W | 211,035 | 0.15 | 0.66 |  | 8.0 | 405,419 | 6.1\% | 6.0\% | 0.9\% | 1.7\% | .7.6\% | 0.7\% |
| UK(Eng\&Wales) | 15W | 23,461 | 0.20 | 0.26 |  | 8.0 | 122,187 | 2.0\% | -3.8\% | 0.3\% | 0.7\% | -2.3\% | 0.2\% |
| UK(N Ireland) 1 | 15W | 31,038 | 0.10 | 0.58 |  | 8.0 | 64,412 | 0.9\% | -1.1\% | 0.1\% | 0.2\% | -1.2\% | 0.1\% |
| UK(N Ireland) 2 | 15W | 10,826 | 0.10 | 0.58 |  | 8.0 | 22,467 | 0.3\% | -0.4\% | 0.1\% | 0.1\% | -0.4\% | 0.0\% |
| UK(Scotland) E | 15W | 21,328 | 0.10 | 0.17 | 34,666 | 7.5 | 187,622 | 2.1\% | -6.3\% | 0.4\% | 0.4\% | -2.8\% | 0.3\% |
| UK(Scotland) W | 15W | 5,568 | 0.20 | 0.11 |  | 8.0 | 68,543 | 1.1\% | -3.7\% | 0.2\% | 0.4\% | -1.3\% | 0.1\% |
| Greenland | 15W | 0 | 0.00 | 1.00 |  | 7.5 | 0 | - | - | - | - | - | - |
| Total Southern Area - 1SW: |  |  |  |  |  |  | 886,179 |  |  |  |  |  |  |
| Southern European Stock Complex - MSW |  |  |  |  |  |  |  |  |  |  |  |  |  |
| France | MSW | 1,277 | 0.00 | 0.30 |  | 17.0 | 5,045 | 0.1\% | -0.3\% | 0.0\% | 0.0\% | -0.2\% | 0.0\% |
| Ireland | MSW | 17,185 | 0.15 | 0.51 |  | 17.0 | 46,910 | 1.5\% | -1.8\% | 0.2\% | 0.4\% | -1.9\% | 0.4\% |
| UK(Eng\&Wales) | MSW | 7,008 | 0.20 | 0.22 |  | 18.0 | 47,671 | 1.6\% | -3.5\% | 0.2\% | 0.6\% | -1.9\% | 0.4\% |
| UK(N Ireland) 1 | MSW | 1,634 | 0.10 | 0.33 |  | 17.0 | 6,621 | 0.2\% | -0.4\% | 0.0\% | 0.0\% | -0.3\% | 0.1\% |
| UK(N Ireland) 2 | MSW | 570 | 0.10 | 0.33 |  | 17.0 | 2,310 | 0.1\% | -0.1\% | 0.0\% | 0.0\% | -0.1\% | 0.0\% |
| UK(Scotland) E | MSW | 25,287 | 0.10 | 0.17 | 34,666 | 17.5 | 238,178 | 5.8\% | -17.3\% | 1.1\% | 1.1\% | -7.8\% | 2.0\% |
| UK(Scotland) W | MSW | 4,487 | 0.20 | 0.11 |  | 17.0 | 60,437 | 2.0\% | 6.8\% | 0.3\% | 0.8\% | -2.4\% | 0.5\% |
| Greenland | MSW | 8,171 | 0.40 | 1.00 |  | 9.0 | 14,901 | - | - | - | - | - | - |
| Total Southern Area - MSW: |  |  |  |  |  |  | 422,074 |  |  |  |  |  |  |

M
0.01

Table 3.7.3.1 Conservation limit options for NEAC stock groups estimated from national lagged lagged egg deposition model and from of river specific values (where available)


Table 3.9.1. Cruises with surface trawling (flotation on trawl wings) in 2000, captures of post-smolts and adult salmon and smolt catch per unit of effort (trawl hours, CPUE)

| $\begin{aligned} & \hline \text { Year } \\ & \text { and } \\ & \text { Cruise } \end{aligned}$ | Gear | Dates | Total number of surface hauls | \% hauls with postsmolt captures | Number of post-smolts captured | Number of salmon captured | CPUE | Area |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000- $1^{2}$ | Harstad ${ }^{\text {C }}$; Fish lift | 06-28.05 | 50 | n.a | n.a. | n.a. | n.a. | Selected fjords SW-Norway (Salmon lice investigatons) |
| 2000- $2^{2}$ | Firkløver trawl ${ }^{\mathbf{B}}$; <br> Fish lift | 10-31.05 | 93 | 42 | 291 | 7 | 2.2 | Norwegian coastal current, Fjords SW - Mid Norway. |
| 2000- $3^{2}$ | Åkra trawl ${ }^{\text {A }}$; Fish lift | 10-20.06 | 14 | 64 | 268 | 6 | 9 | Norwegian Sea, Voering plateau |
| $\begin{aligned} & 2000-4 \\ & 2000-5^{2} \end{aligned}$ | Åkra trawl ${ }^{\mathbf{A}}$ <br> Firkløver trawl ${ }^{\mathbf{B}}$; <br> Fish lift | $\begin{aligned} & 24-28.06 \\ & 28.06-24.07 \end{aligned}$ | 2 106 | 0 30 | 0 202 | 0 13 | $\begin{gathered} 0 \\ 1.22 \end{gathered}$ | The Halten Bank, Norwegian Sea Selected fjords N-Norway and S-Barents Sea (special salmon cruise) |
| $\begin{aligned} & 2000-6 \\ & 2000-7 \end{aligned}$ | Åkra trawl ${ }^{\mathbf{A}}$ <br> Harstad $25 \times 25 m$ <br> float- trawl | $\begin{aligned} & 21.07-16.08 \\ & 17.08-07.09 \end{aligned}$ | 26 3 | 12 n.a | 5 0 | ${ }_{1}{ }^{1}$ | $\begin{gathered} 0.38 \\ 0 \end{gathered}$ | Norwegian Sea Western Barents Sea |
| TOTAL | 2000 |  | 294 |  | 766 | 38 |  |  |

${ }^{1}$ The salmon was captured in a sub-surface trawl (haul)
${ }^{2}$ Cruises dedicated to salmon investigations
${ }^{\text {A }}$ Dimensions of the $\AA$ Ara trawl opening $25 \times 25 \mathrm{~m}$
${ }^{\text {B }}$ Dimensions of the Firkløver trawl opening $18.5 \times 18.5 \mathrm{~m}$
${ }^{C}$ Dimensions of the Harstad float trawl opening $14 \times 14 \mathrm{~m}$

Figure 3.3.3.1 Percent changes in gear units licened or authorised for net fishery in the NEAC are since 1996 (5-year period).


Figure 3.3.3.2 Percent changes in number of rod days, or rods licened or authorised in the NEAC are since 1996 (5-year period).


[^8]Figure 3.3.4.1 Nominal catches of salmon in the NEAC area in year 2000 relative to mean catches in the preceeding 10-year and 5-year period. Countries are ordered after size of change.


Figure 3.3.6.1. The proportions of 1SW salmon in the NEAC catches in 2000 relative to previous indices



Fig. 3.4.1.2. Conservation limit (CL) attainment (egg deposition/conservation limit, \%) for index
rivers having a mean egg deposition below (Type A), around (Type B) or above (Types C\&D) the CL over the 10 previous years the CL over the 10 previous years.




Figure 3.6.4.1 Estimated recruitment (PFA) in the NEAC Area, 1970-2000
a) Maturing 1SW recruits (potential 1SW returns)
(Recruits in Year N become spawners in Year N)

b) Non-maturing 1SW recruits (potential MSW returns)
(Recruits in Year N become spawners in Year N+1)


Figure 3.6.4.2 Estimated spawning escapement in the NEAC Area, 1970-2000
a) 1SW spawners (and 95\% confidence limits)

b) MSW spawners (and 95\% confidenance limits)


Figure 3.6.4.3 Estimated recruitment (PFA) and Spawning Escapement Reserve (SER) for maturing and non-maturing salmon in Northern Europe, 1971-2000.
a) Maturing 1 SW recruits (potential 1 SW returns)
(Recruits in Year $N$ become spawners in Year $N$ )

b) Non-maturing 1 SW recruits (potential MSW returns)
(Recruits in Year $N$ become spawners in Year $N+1$ )


Figure 3.6.4.4 Estimated spawning escapement of maturing and nonmaturing salmon in Northern Europe, 1971-2000.
a) 1SW spawners (and 95\% confidence limits)

b) MSW spawners (and 95\% confidence limits)


Figure 3.6.4.5 Estimated recruitment (PFA) and Spawning Escapement Reserve (SER) for maturing and non-maturing salmon in Southern Europe, 1971-2000.
a) Maturing 1SW recruits (PFA) (potential 1SW returns)
(Recruits in Year N become spawners in Year N )

b) Non-maturing 1SW recruits (potential MSW returns)
(Recruits in Year N become spawners in Year $\mathrm{N}+1$ )


Figure 3.6.4.6 Estimated spawning escapement of maturing and nonmaturing salmon in Southern Europe, 1971-2000.
a) 1SW spawners (and 95\% confidence limits)

b) MSW spawners (and 95\% confidence limits)


Figure 3.7.2.1 Demonstration of the simple model for establishing conservation limits

Top right: Simulated stock recruitment data with little variability
Top left: Model output for low variability data - residual sum of squares against change points

Bottom right: Simulated data with large variability
Bottom left: Model output for high variability data - Residual sum of squares against change points



Figure 3.9.1 Post-smolt and salmon captures in surface trawl hauls in 2000 (legends in figure). Area of captures at special salmon survey in June, and sites of capture of micro-tagged and Carlin-tagged fish encircled. Shaded circle further north indicates area of captures of micro-tagged fish in July/ August 1995-97. EEZs of countries bordering the NE-Atlantic are drawn as dashed lines.


Figure 3.9.2 Surface trawl captures of post-smolts at the coast in the Norwegian Coastal Current (drawn as grey arrows). Stars mark trawl hauls with no salmon, numbers mark post-smolts in the catch


Figure 3.9.3 Surface trawl surveys in four fjords in northern Norway 28 June - 24.July. Catch legends as in figure 3.9.2. The coastal current is presented as a grey arrow, while the warmer more saline Atlantic current is presented as undulating lines

### 4.1 Description of Fisheries

### 4.1.1 Gear and effort

## Canada

The 23 areas for which the Department and Fisheries and Oceans (DFO) manages the salmon fisheries are called Salmon Fishing Areas (SFAs); for Québec, the management is delegated to the Société de la Faune et des Parcs du Québec and the fishing areas are designated by Q1 through Q11 (Figure 4.1.1.1). Harvest (fish which are killed and retained) and catches (including harvests and fish caught-and-released in recreational fisheries) are categorized in two size groups: small and large. Small salmon in the recreational fisheries refer to salmon less than 63 cm fork length, whereas in commercial fisheries, it refers to salmon less than 2.7 kg whole weight. Large salmon in recreational fisheries are greater than or equal to 63 cm fork length and in commercial fisheries refer to salmon greater than or equal to 2.7 kg whole weight.

Three user groups exploited salmon in Canada in 2000: Native peoples, residents fishing for food in Labrador and recreational fishers. Commercial fisheries in Québec which operated in 1999 in zone Q9 were closed and licenses bought back in 2000. Commercial quotas normally fished by Native peoples in Ungava Bay (zone Q11) remained closed. Hence there were no commercial fisheries in Canada in 2000.

The following management measures were in effect in 2000:

Native peoples' food fisheries: In Québec, Native peoples' food fisheries took place subject to agreements or through permits issued to the bands. There are 10 bands with subsistence fisheries in addition to the fishing activities of the Inuit in Ungava (Q11), who fished in estuaries or within rivers. The permits generally stipulate gear, season and catch limits. Catches for subsistence fisheries have to be reported collectively by each native user group. However, if reports are not available, the catches are estimated. In the Maritimes and Newfoundland (SFAs 1 to 23), food fishery harvest agreements were signed with several Native peoples groups (mostly First Nations) in 2000. The signed agreements often included allocations of small and large salmon. Harvests which occurred both within and outside agreements were obtained directly from the Native peoples. Harvest by Native peoples with recreational or commercial licenses are reported under the recreational and commercial harvest categories.

Residents food fisheries in Labrador: In the Lake Melville (SFA 1) and the coastal southern Labrador (SFA 2) areas, DFO allowed a food fishery for local residents. Residents who requested a license were permitted to retain a maximum of four (4) salmon of any size while fishing for trout and charr; 4 salmon tags accompanied each license. The license restricted the fishing gear to a gillnet of 15 fathoms ( 27.4 m ) and $3.5 \mathrm{in}(89 \mathrm{~mm})$ mesh. The seasons were June 15-July 2 and July 24-August 19 in SFA 1 and July 15-August 31 in SFA 2. All licensees were to complete logbooks. A total of 970 licenses were issued.

Recreational fisheries: Recreational fisheries management in 2000 varied by area (Figure 4.1.1.2). Except in Québec and Labrador (SFA 1 and 2), only small salmon could be retained in the recreational fisheries.

The seasonal bag limits in the recreational fishery remained at eight small salmon in New Brunswick and in Nova Scotia. In SFA 16 and in Nepisiquit River (SFA 15) of New Brunswick, the small salmon daily retention limit remained at one fish. In the remainder of SFA 15 and in Nova Scotia (SFA 18), the daily retention limits were two small salmon. The maximum daily catch limit was four fish daily. In SFA 17 (PEI), the season and daily bag limits were 7 and 1 respectively. Catch-and-release fishing only for all sizes of Atlantic salmon was in effect in SFA 19 of Nova Scotia. SFAs 20-23 of Nova Scotia and New Brunswick were closed to all salmon angling, except for five acidimpacted rivers on the Atlantic coast of Nova Scotia, where retention of small salmon, mostly of hatchery origin, was allowed.

For insular Newfoundland (SFAs 3 to 14A) and the Strait of Belle Isle shore of Labrador (SFA 14B), the second year of a three year management plan was continued for the recreational fishery which allowed differing seasonal retention limits based on the status of the salmon stocks in the rivers. Retention limits ranged from a seasonal limit of 6 fish on Class I rivers, to no retention and catch-and-release only on Class IV rivers (five rivers in 2000). Some rivers were closed to all angling and were not assigned a class number. The river classification scheme rated individual rivers as Class I (highest) to Class IV (lowest) according to their ability to sustain angling activities as follows:

Class I - large rivers with a seasonal bag limit of 6 fish,

Class II - smaller rivers with a season bag limit of 4 fish,

Class III - rivers with a season bag limit of 2 fish,
Class IV - rivers with catch and release only.

Special class - with various management plans.
In SFAs 1 and 2 of Labrador, there was a seasonal limit of four fish, only one of which could be a large salmon.
In Québec, management rules were set before the season opening as a way to reach conservation limits on each river. Fishing licenses allowed a total landing of 7 salmon for the year. The northern zones (Q8, Q9 and Q11) include 44 rivers which were managed mainly on a zonal basis. Sport fishing was permitted on all rivers except one and retention of both small and large salmon was allowed throughout the northern zones. The daily limit was three fish in Q9, two in Q8 and one in zone Q11. Release of large salmon occurred mainly on a voluntary basis. The 74 rivers of the southern zones were managed river by river. Fishing was not allowed on nine rivers, retention of small salmon only was in force on 34 rivers and retention of small and large was allowed on 31 rivers. On these rivers, fishing for the day would end if the first fish caught was a large salmon. If the first fish was a small salmon, then fishing could continue on most rivers until the second fish, small or large was caught. Seven additional rivers were restricted to retention of small salmon only after mid-season reviews as an insufficient numbers of spawners were detected by this time in the rivers. (Figure 4.1.1.2).

## USA

There was no fishery for Atlantic salmon in the USA in 2000; in 1999, hook-and-release fishing only was permitted in some rivers.

## France (Islands of Saint-Pierre and Miquelon)

For the Saint-Pierre and Miquelon fisheries in 2000, there were 8 professional and 35 recreational gillnet licenses issued. The number of professional fishermen has increased by one license from 1999 and the number of recreational licenses decreased by five licenses since 1999.

| Year | Number of <br> Professional <br> Fishermen | Number of <br> Recreational <br> Licenses |
| :---: | :---: | :---: |
| $\mathbf{1 9 9 5}$ | 12 | 42 |
| $\mathbf{1 9 9 6}$ | 12 | 42 |
| $\mathbf{1 9 9 7}$ | 6 | 36 |
| $\mathbf{1 9 9 8}$ | 9 | 42 |
| $\mathbf{1 9 9 9}$ | 7 | 40 |
| $\mathbf{2 0 0 0}$ | 8 | 35 |

### 4.1.2 Catch and catch per unit effort (CPUE)

## Canada

The provisional harvest of salmon in 2000 by all users was 150 t , similar to the 1999 harvest of 152 t (Table 2.1.1.1; Figure 4.1.2.1). The 2000 harvest was 50,108 small salmon and 11,458 large salmon, about the same as the 1999 harvests for both (Table 4.1.2.1). The dramatic decline in harvested tonnage since 1988 is in large part the result of the reductions in commercial fisheries effort, the closure of the insular Newfoundland commercial fishery in 1992, the closure of the Labrador commercial fishery in 1998 and the closure of theQuébec commercial fishery in 2000 (Figure 4.1.2.1). These reductions were introduced as a result of declining abundance of salmon.

The 2000 harvest of small and large salmon, by number, was divided among the three user groups in different proportions depending on the province and the fish-size group exploited (Table 4.1.2.1). Newfoundland reported the
largest proportion of the total harvest of small salmon and Québec reported the greatest share of the large salmon harvest. Recreational fisheries exploited the greatest number of small salmon in each province, accounting for $79 \%$ of the total small salmon harvests in eastern Canada. Unlike years previous to 1999 when commercial fisheries took the largest share of large salmon, food fisheries (including the Labrador resident food fishery) accounted for the largest share in 2000 ( $60 \%$ by number).

Native peoples' food fisheries. Harvests in 2000 (by weight) were up $10 \%$ from 1999 and $18 \%$ above the previous 5year average harvest. In some cases, particularly in the Maritime provinces, Native peoples' food fisheries harvests in 2000 were less than the allocations.

| Native peoples' food fisheries |  |  |  |
| :--- | :---: | :---: | :---: |
|  |  | \% large |  |
| Year | Harvest (t) | by weight | by number |
| $\mathbf{1 9 9 0}$ | 31.9 | 78 |  |
| $\mathbf{1 9 9 1}$ | 29.1 | 87 |  |
| $\mathbf{1 9 9 2}$ | 34.2 | 83 |  |
| $\mathbf{1 9 9 3}$ | 42.6 | 83 |  |
| $\mathbf{1 9 9 4}$ | 41.7 | 83 | 58 |
| $\mathbf{1 9 9 5}$ | 32.8 | 82 | 56 |
| $\mathbf{1 9 9 6}$ | 47.9 | 87 | 65 |
| $\mathbf{1 9 9 7}$ | 39.4 | 91 | 74 |
| $\mathbf{1 9 9 8}$ | 47.9 | 83 | 63 |
| $\mathbf{1 9 9 9}$ | 45.9 | 73 | 49 |
| $\mathbf{2 0 0 0}$ | 50.3 | 69 | 43 |

Residents Fishing for Food in Labrador: A total of 633 logbooks have been processed for this fishery. The estimated catch for the entire fishery in 2000 was 5.6 t , about 2,300 fish ( $79 \%$ small salmon by number).

Recreational fisheries: Harvest in recreational fisheries in 2000 totaled 44,412 small and large salmon, $20 \%$ below the previous 5 -year average and $8 \%$ below the 1999 harvest level (Figure 4.1.2.2). The small salmon harvest of 39,785 fish was a decrease of $19 \%$ from the previous 5 -year mean. The large salmon harvest of 4,627 fish was a $32 \%$ decline from the previous five-year mean. Small and large salmon harvests were down $9 \%$ and $4 \%$ from 1999 , respectively. The small salmon size group has contributed $87 \%$ on average of the total harvests since the imposition of catch-and-release recreational fisheries in the Maritimes and insular Newfoundland (SFA 3 to 14B, 15 to 23) in 1984 (Figure 4.1.2.2).

Recreational catches (including retained and released fish) of small salmon in 2000 were similar or above the 1984 to 1991 mean in some fishing areas of Québec (Q1,Q2,Q3,Q5), and the north-east coast and northern peninsula of Newfoundland and throughout Labrador (Figure 4.1.2.3). Small salmon catches were among the lowest observed in the majority of the Maritimes. Large salmon catches were among the lowest observed throughout mainland Canada but were among the highest in the west coast of Newfoundland, (SFA 13, 14A) and Labrador (SFAs 1,2, and 14B).

In 1984, anglers were required to release all large salmon in the Maritime provinces and insular Newfoundland. Changes in the management of the recreational fisheries since 1984 have compromised the use of angling catches as indices of abundance. Therefore, the interpretation of trends in abundance relies mostly on rivers where returns have been estimated or completely enumerated. Caught-and-released fish are not considered equivalent to retained fish and their inclusion in catch statistics further compromises the reliability of interpretation of trends. In more recent years, anglers have been required to release all salmon on some rivers for conservation reasons and, on others, they are voluntarily releasing angled fish. In addition, numerous areas in the Maritimes Region in 2000 were closed to retention of all sizes of salmon (Figure 4.1.1.2).

Hook-and-released salmon fisheries: In 2000, about 49,700 salmon (about 20,700 large and 29,000 small) were caught and released (Table 4.1.2.2), representing about $51 \%$ of the total number caught, including retained fish. This was a $1 \%$ increase from the number released in 1999. Most of the fish released were in New Brunswick (44\%), followed by Newfoundland (43\%), Québec (9\%), Nova Scotia (3\%) and Prince Edward Island ( $0.3 \%$ ). Expressed as a proportion of the fish caught, that is, the sum of the retained and released fish, Nova Scotia released the highest percentage ( $84 \%$ ), followed by New Brunswick (57\%), Newfoundland (56\%), Prince Edward Island (49\%) and Québec (31\%).

Commercial fisheries: All commercial fisheries for Atlantic salmon were closed in Canada in 2000 and the catch therefore was 0 . Catches have decreased from a peak in 1980 of almost $2,500 \mathrm{t}$ to 0 currently as a result of effort reductions, low abundance of stocks and closures of fisheries (Figure 4.1.2.4).

Unreported catches: Canada's unreported catch estimate for 2000 is about 124 t , compared to 133 t in 1999. Estimates were included for all provinces (but not for all areas within the provinces) and were provided mainly by scientific staff. In the many cases where enforcement staff did not respond to requests for estimates, values previously provided by them were assumed for 2000. An exception to this was Labrador where the unreported catch estimate for 1999 was 6.3 t . No new estimate was provided for 2000 and it was determined to be inappropriate to use the 1999 estimate as most of the unreported catch in 1999 was ascribed to fisheries that are now included in the reported catch. In all areas, most unreported catch arises from illegal fishing or illegal retention of bycatch of salmon.

By stock groupings used for Canadian stocks throughout the report, the unreported catch estimates for 2000 were:

| Stock Area | Unreported Catch $(\mathrm{t})$ |
| :--- | :---: |
| Labrador | No estimate |
| Newfoundland | 65 |
| Gulf | 39 |
| Scotia-Fundy | $<1$ |
| Québec | 19 |
| Total | 124 |

## USA

All fisheries (commercial and recreational) for sea-run Atlantic salmon within the USA are now closed, including rivers previously open to catch-and-release fishing. Thus, there was no harvest of sea-run Atlantic salmon in the USA in 2000. Incidental catch was four Atlantic salmon, caught and released as the result of angling for other species (three on the Penobscot River and one on the Narraguagus River). As in 1999, unreported catches in the USA were estimated to be 0 t .

## France (Islands of Saint-Pierre and Miquelon)

The harvest in 2000 was reported to be 2.3 t , the same as in 1998 and 1999, split about equally between professional and recreational fishermen in 2000 (Table 2.1.1.1). There was no estimate made of unreported catch for 2000.

### 4.1.3 Origin and composition of catches

In the past, salmon from both Canada and USA have been taken in the commercial fisheries of eastern Canada. These fisheries have since been closed. The remaining Native Peoples' and resident food fisheries that exist in Labrador may intercept some salmon from other areas of North America although there are no reports of tagged fish being captured there in 2000. The fisheries of Saint Pierre and Miquelon catch salmon of both Canadian and US origin. Little sampling occurs in these remaining fisheries and often the best information on stock origins comes from sampling of the returns to rivers, rather than the fisheries.

Fish designated as being of wild origin are defined as the progeny of fish where mate selection occurred naturally (eggs not stripped and fertilised artificially) and whose life cycle is completed in the natural environment (ICES 1997/Assess:10). Hatchery-origin fish, designated as fish introduced into the rivers regardless of life stage, were identified on the basis of the presence of an adipose clip, from fin deformations, and/or from scale characteristics. Not all hatchery fish could be identified as such in the returns because of stocking in the early life stages. Aquaculture escapees were identified from hatchery fish on the basis of fin erosion (especially of the tail) and from scale characteristics.

The returns to the majority of the rivers in Newfoundland and to most rivers of the Gulf of St. Lawrence and Québec were comprised exclusively of wild salmon (Figure 4.1.3.1). Hatchery origin salmon made up varying proportions of the total returns and were most abundant in the rivers of the Bay of Fundy, the Atlantic coast of Nova Scotia and the USA. Aquaculture escapees were present in the returns to four rivers of the Bay of Fundy and the coast of Maine (St. Croix, Magaguadavic, Union and Dennys) in 2000.

Aquaculture production of Atlantic salmon in eastern Canada has increased annually, exceeding 10,000 t in 1992 and rising to almost $32,000 \mathrm{t}$ in 2000 (Table 2.2.1.1). Escapes of Atlantic salmon have occurred annually. In 1994, escapes of Atlantic salmon in the Bay of Fundy area were estimated at 20,000 to 40,000 salmon, an amount greater than the total returns of wild and hatchery origin salmon (both small and large) ( 13,000 to 21,000 fish) to the entire Bay of Fundy and Atlantic coast of Nova Scotia area (SFA 19 to 23) in the same year. The documented minimum numbers of farmed salmon that escaped in 1999 and 2000 from the North American east coast industry (Canada and USA combined) were 50,000 and 118,000 respectively (see Section 2.4.2).

The proportion of the run that are aquaculture escapees has been high (greater than $50 \%$ ) since 1994 in the Magaguadavic River (SFA 23; Table 4.1.3.1) which is in close proximity to the centre of the aquaculture production area (Figure 4.1.3.2). Escaped fish were not observed between 1983 and 1988. Since 1992, escaped fish have comprised between $33 \%$ and $90 \%$ of the total counts at the fishway. Aquaculture escapees comprised between $13 \%$ and $64 \%$ of the total run of salmon to the St. Croix River during 1994 to 2000 (Table 4.1.3.1). In addition to the St. Croix River, three other rivers in Maine were monitored for the occurrence of farmed salmon in 2000, the Union, Dennys and Narraguagus. Percentages of returns which were of farmed origin were 60,94 and $0 \%$, respectively.

### 4.1.4 Exploitation rates in Canadian and USA fisheries

## Canada

In Newfoundland, exploitation rates were available for seven rivers in 2000. For those rivers with retention of small salmon, exploitation rates ranged from $6 \%$ to $28 \%$. In Labrador, exploitation rates were available for two rivers; exploitation rates for small salmon range from $2 \%$ to $9 \%$ and 0 to $2 \%$ for large salmon.

In Québec, exploitation rates were estimated using mid-point estimates of returns and recreational landings. Exploitation rates for small salmon were $18 \%$ and for large salmon, $10 \%$.

In previous years, overall Canadian exploitation rates were calculated as the harvest of salmon divided by the estimated returns to North America. No estimates of returns to Labrador are possible for 1998, 1999 and 2000, as there was no commercial fishery and there was insufficient information collected on freshwater escapements to extrapolate to other Labrador rivers. For this reason, exploitation rates cannot be calculated for 1998-2000. Harvests in 2000 of 50,108 small and 11,458 large salmon were less than those of 1997, substantially in the case of large salmon. Exploitation rates in 1997 were estimated to be between 0.14 and 0.26 for small and 0.15 and 0.25 for large salmon.

## USA

There was no exploitation of USA salmon in homewaters and no salmon of USA origin were reported in Canadian fisheries in 2000.

### 4.2 Status of Stocks in the North American Commission Area

There are approximately 550 Atlantic salmon rivers in eastern Canada and 21 rivers in eastern USA each of which could contain at least one population of salmon. Assessments are prepared for a limited number of specific rivers, because they compose significant fractions of the salmon resource or are indicators of patterns within a region, or because of the demands by user groups, or as a result of requests for biological advice from fisheries management. The status is evaluated by examining trends in returns and escapement relative to the conservation requirements.

### 4.2.1 Measures of abundance in monitored rivers

## Canada

The returns represent the size of the population before any in-river and estuarine removals. Spawning escapement is determined by subtracting all the known removals, including food fisheries, recreational harvests, broodstock collections, and scientific samples from the total returns.

A total of 72 rivers were assessed in eastern Canada in 2000. Estimates of total returns of small and large salmon were obtained using various techniques: 37 were derived from counts at fishways and counting fences; 4 were obtained using mark and recapture experiments; 24 using visual counts by snorkeling or from shore; and 7 from angling catches, and redd counts.

2000 compared to 1999 adult returns: Of the 72 stocks for which returns of salmon were determined in 2000, comparable data were available for 68 of these in 1999. For 51 of these rivers, returns were estimated by small salmon and large salmon size groups separately in both years (Table 4.2.1.1). For both size groups combined, returns in 2000 were less than $50 \%$ of the 1999 returns in seven of the rivers assessed ( $10 \%$ ), between $50 \%$ and $90 \%$ of 1999 returns in $28(41 \%)$ of the rivers and were $90 \%$ or greater than 1999 returns in $33(49 \%)$ of the rivers. The Newfoundland rivers showed the highest number of improvements in returns.

Large salmon returns in 2000 decreased from 1999 in rivers throughout the Maritime provinces (63\%) and in Québec ( $85 \%$ ). Lower proportions of the rivers were down or improved in Newfoundland (52\%) (Table 4.2.1.1; Figure 4.2.1.1). In most of the rivers of Newfoundland, except for rivers of the south-west coast (SFA 13), large salmon are mostly repeatspawning 1SW fish.

Small salmon returns in 2000 relative to 1999 were generally reduced throughout eastern Canada in the majority of their monitored rivers except in Québec (23\%) (Table 4.2.1.1, Figure 4.2.1.1). Returns were similar to or improved ( $>90 \%$ in 2000 relative to 1999 ) in about half ( $53 \%$ ) of the assessed rivers but decrease to $45 \%$ excluding Québec's rivers.

1985-2000 patterns of adult returns: Annual returns of salmon by size group are available for 26 rivers in eastern Canada since 1985. These returns do not account for commercial fisheries removals in Newfoundland, Labrador, Québec and Greenland and in some rivers include returns from hatchery stocking. Peak return years differed for regions within eastern Canada (Figure 4.2.1.2). The returns during the Newfoundland commercial fishery moratorium years (1992 to 2000) for all areas except Newfoundland are lower than returns in 1986 to 1988 when there were commercial fisheries in Newfoundland, Labrador, Québec and Greenland harvesting mainland Canada origin salmon. The total returns to seven Newfoundland rivers doubled during 1993 to 2000 from the low levels observed during 1989 to 1991 (Figure 4.2.1.2).

The returns for 2000 of large salmon in all areas except Newfoundland were among the lowest observed during the last 15 years and decreased in all regions except for the Gulf (Table 4.2.1.1, Figure 4.2.1.2). The returns of large salmon in 2000 were the lowest of the time series for the Nova Scotia and Bay of Fundy with a decrease of $68 \%$. Returns of large salmon to ten rivers of Québec in 2000 were the second lowest since 1985. Returns of small salmon in Québec and Gulf rivers in 2000 increased from 1999. Returns to the rivers of the Atlantic coast (Nova Scotia and Bay of Fundy) and Newfoundland decreased for small salmon by 8 and $35 \%$ respectively.

Smolt and juvenile abundance: Counts of smolts provide direct measurements of the outputs from the freshwater habitat. Previous reports have documented the high annual variability in the annual smolt output: in tributaries, smolt output can vary by five times but in the counts for entire rivers, annual smolt output has generally varied in magnitude by a factor of two. Wild smolt production has been estimated in 11 rivers of eastern Canada, although only nine rivers have several years of data (Figure 4.2.1.3). In other rivers, juvenile abundance surveys have been conducted.

In 2000, smolt production from the two monitored rivers in Québec was among the lowest of the time series and about half of the 1990-95 average (Figure 4.2.1.4). In Newfoundland, smolt production in 2000 remained at or just above the 1990 to 1995 average but lower than recent years in three of the five rivers (Figure 4.2.1.4).

Juvenile salmon abundance has been monitored annually since 1971 in the Miramichi (SFA 16) and Restigouche (SFA 15) rivers and for shorter and variable time periods in other rivers (Figure 4.2.1.5). In the rivers of the southern Gulf (SFAs 15,16 and 18), densities of young-of-the-year (fry) and parr (juveniles of one or more years old) have increased since 1985 in response to increased spawning escapements (Figure 4.2.1.5). Densities of parr in 2000 increased to record values in the Restigouche River. In the Miramichi River, both fry and parr densities remained high in 2000 but dropped to near-mean values for 1985-1999. High densities of juveniles have also been reported from Nova Scotia rivers along the Gulf of St. Lawrence (SFA 18) and in several Cape Breton Island streams (SFA 19).

Rivers of SFAs 20 and 21 along the Atlantic coast of Nova Scotia are generally organic stained, of lower productivity, and, when combined with acid precipitation, can result in acidic conditions toxic to salmon. Prognoses for salmon populations in 47 of 65 of these rivers indicate that 40 populations are now likely to be extirpated. In the low-acidified St. Mary's River, fry (age $0^{+}$) densities remain at moderate abundance while total parr (age- $1^{+}$and older) densities are the lowest since 1993 (Figure 4.2.1.5). Juvenile densities in the outer Bay of Fundy (SFA 23) Saint John River and its tributaries, the Nashwaak, Hammond and Kennebecasis rivers, have declined since 1984 as a result of reduced spawning escapement.

Juvenile densities from inner Bay of Fundy rivers (SFA22 and a portion of SFA 23) are critically low and contribute to evidence for possible listing of these stocks pending the passage of the Species at Risk Act. In 2000, densities of parr in 33 electrofishing stations of the Stewiacke River were at a record low (Figure 4.2.1.5); no fry were found.

It is not possible to measure the total smolt production from the rivers of Atlantic Canada for any given year. However, juvenile abundance indices were considered as surrogates of smolt production from eastern Canada. To allow for the combined analysis of smolt counts and juvenile abundance surveys from all the rivers, the individual river surveys were standardized to a common period, 1995 to 1998.

$$
\begin{aligned}
& \begin{array}{l}
\operatorname{Ind}_{i j} \\
{\text { where } \operatorname{Ind}_{i j}}=\quad=\quad \text { Abund }_{i j} / \text { Average }_{j} \\
\text { Adjusted index of juvenile or smolt abundance } \quad \text { for year } i \text { and river } j
\end{array} \\
& \text { Abund }_{i j}=\quad \begin{array}{l}
\text { Measured abundance of juvenile or smolts } \\
\text { for year } i \text { and river } j
\end{array} \\
& \text { Average }_{j}=\quad \text { Average abundance for years } 1995 \text { to } 1998 \text { in river } j
\end{aligned}
$$

This adjustment places all the rivers on a common scale and provides a measure of the temporal variability in the smolt and juvenile measures. Juvenile measures were age 1 and older parr and were lagged forward one year to correspond to the smolt migration year.

The index of smolts from North America was obtained by weighting the annual river indices by the relative proportion of the conservation egg requirements (O'Connell et al. 1997) of the SFA or Zone to the total conservation egg requirements of the zones under consideration (Table 4.2.1.2). An alternative weighting incorporated the relative contribution to the 2SW spawner requirements of the areas or zones within North America. This allows indices of smolt production from all areas of North America to be used but attributes weights to the area indices according to the expected contribution to 2 SW abundance.

The longest time series are from Western Arm Brook (SFA 14A) in Newfoundland and the Miramichi and Restigouche rivers in the Gulf (SFAs 15 and 16). The number of rivers with available data has increased from two in 1971 to greater than 25 rivers since 1995 (Table 4.2.1.3). The proportion of the indexed areas represented by the index rivers has increased from $11 \%$ in 1971 to more than $25 \%$ since 1993 (Table 4.2.1.3).

The relative index of smolt production, weighted by the area-index factor, indicates relative smolt production at three levels since 1971: at about one-third the 1995 to 1998 average between 1971 and 1979 , at about $60 \%$ of the average during 1980 to 1989 and at about average since 1991 (Figure 4.2.1.6). The relative index for 2SW recruitment (calculated by excluding the Newfoundland areas which do not produce 2SW salmon, SFA 3-12, 14A or by weighting all areas according to the 2 SW spawner requirements by area) suggests an overall similar trend.

Estimates of the relative smolt index in the four geographic areas correspond to the previously documented status of rivers (Figure 4.2.1.7). Smolt production from Newfoundland rivers has approximately doubled over the 1971 to 2000 time period (Figure 4.2.1.7). The Gulf smolt index is at its highest level in the 1990s. TheQuébec smolt index has declined between 1984 and 2000, driven by de la Trinité time series which forQuébec has a large area-index weight (Table 4.2.1.2). The relative index for Scotia-Fundy has essentially remained unchanged.

## USA

The documented returns of Atlantic salmon to rivers in New England in 2000 was 803 fish. Returns of 1SW salmon were 270 compared to 358 in 1998 and 386 in 1999, while MSW returns decreased to 533 . MSW returns were over 1,000 in the two previous years. Total salmon returns to the rivers of New England continued the downward trend that began in the mid-1980s, and were lower than the previous 5 -year and 10-year averages. Documented Atlantic salmon returns to USA rivers since 1967 are shown in Figure 4.2.1.8. These are minimal estimates, since many rivers in Maine do not contain fish counting facilities, and where counting facilities exist, they do not count $100 \%$ of the returns.

The majority of the returns in 2000 were recorded in the rivers of Maine, with the Penobscot River accounting for nearly $67 \%$ of the total New England returns. The Connecticut River adult returns accounted for about $10 \%$ of the New England total and $46 \%$ of the adult returns outside of Maine. Overall, $34 \%$ of the adult returns to New England were 1SW salmon and $66 \%$ were MSW salmon. Most (72\%) of these fish were of hatchery smolt origin. Of the total returns, approximately $28 \%$ were from natural reproduction and stocked fry.

### 4.2.2 Estimates of total abundance by geographic area

For assessment purposes, the following regions were considered: Labrador (SFA 1, 2, \& 14B), Newfoundland (SFA 3-14A), Québec (Q1-Q11), Gulf of St. Lawrence (SFA 15-18), Scotia-Fundy (SFA 19-23) and USA. Returns of 1SW and 2 SW salmon to each region (Tables 4.2.2.1 and 4.2.2.2; Figures 4.2.2.1 and 4.2.2.2; and Appendix 5) were
estimated by updating the methods and variables used by Rago et al. (1993b) and reported in ICES 1993/Assess:10. The returns for both sea-age groups were derived using a variety of methods using data available for individual river systems and management areas. These methods included counts of salmon at monitoring facilities, population estimates from mark-recapture studies, and the application of angling and commercial catch statistics, angling exploitation rates, and measurements of freshwater habitat (Appendix 5). The 2SW component of the MSW returns was determined using the sea-age composition of one or more indicator stocks.

In the context used here "returns" are the number of salmon that returned to the geographic region, including homewater commercial fisheries, except in the case of Newfoundland and Labrador regions where returns do not include commercial fisheries. This was done to avoid double counting of fish when commercial catches in Newfoundland and Labrador are added to returns of all geographic areas in North America to create the PFA of North American salmon.

Labrador: The basis for estimates of 2SW and 1SW salmon returns and spawners for Labrador (SFAs 1, 2 \& 14B) prior to 1998 are catch data from angling and commercial fisheries. Catch and effort data from the angling fishery were collected by DFO enforcement staff in conjunction with angling reports submitted by fish camp operators and processed by DFO Science Branch personnel. In 1997 for SFA 14B, the angling catch statistics were derived from a licence stub system similar to insular Newfoundland while in SFAs $1 \& 2$ the camp statistics data were used. Commercial catch data were collected by DFO enforcement staff from fish plant landing slips and processed by DFO Statistics and Informatics Branch personnel.

In 1998-2000, there was no commercial fishery in Labrador and although counting projects took place in 2000 on two small Labrador rivers, out of about 100 salmon rivers that exist, it is not possible to extrapolate from these rivers to unsurveyed ones. For Labrador, returns were previously estimated from commercial catches and exploitation rates. As there was no commercial fishery since 1998, it was not possible to estimate the returns or spawners to Labrador for these years.

Newfoundland: The estimates of 1SW and 2SW returns and spawners for insular Newfoundland (SFAs 3-12 \& 14A) are updated for the entire time series. Prior to 1999, they are derived from exploitation rates estimated from rivers with counting facilities which are subsequently applied to angling catches of small salmon, adjusted for the proportions of large:small salmon at counting facilities, and finally the proportion of large salmon that are 2 SW . Exploitation rates for small salmon (retained only) were calculated by dividing the total count and the catch (retained) from rivers with enumeration facilities. In 1997, for SFAs 3-14A, angling catch data was derived from the license stub return system (O'Connell et al. 1997) while in previous years angling catch data was collected by DFO Fishery Officers and Guardian staff. For SFA 13, returns and spawners come from four assessment facilities expanded to the entire drainage area based on their proportionate contribution. The time series of ratios of large:small salmon were updated to include counts for some rivers not previously available. In last years report, a precision index was developed for Newfoundland based on known counts at enumeration facilities. It was noted that in some instances the index for the minimum value was higher than one, indicating that for some years, more salmon were counted than estimated. In order to correct this, exploitation rates were revised based on a weighted average of those rates available for assessed rivers. The new exploitation rates are:

| Year | $95^{\text {th }}$. C.I. | $5^{\text {th }}$. C.I. |
| :---: | :---: | :---: |
| 1995 | 0.1460 | 0.0674 |
| 1996 | 0.1500 | 0.0624 |
| 1997 | 0.1906 | 0.1149 |
| 1998 | 0.1360 | 0.0758 |

The dramatic changes to estimates of returns of both 1SW and 2SW salmon that occurred in 1995 and 1996 were due to lower exploitation rates for the three largest rivers in Newfoundland compared to the smaller ones (Figure 4.2.2.3). Therefore, when averages were weighted to the size of the returns to the rivers and were used instead of unweighted, the lower exploitation rate resulted in higher population estimates. The weighted averages were considered to be more appropriate than unweighted ones as the larger rivers have higher returns which would not have been as well estimated with the higher unweighted exploitation rates.

Beginning in 1999, the method used in previous years was modified to take into consideration the changes implemented in the 1999-2001 Salmon Management Plan. The Management Plan introduced, for the first time, a river classification scheme with different season limits for each of classes I-IV and, in addition, some other rivers were placed in a special class with a different management plan for each river. Since the intent of the Management Plan was to alter exploitation for rivers in the various classes, it would be better to model the estimation procedure for returns and spawners individually for each of them. However, there are too few rivers in each class to do so, there only being ten
rivers available in 1999 and seven in 2000 with which to estimate parameter values for three river classes separately. The $95^{\text {th }}$ confidence intervals of bootstrap estimates of unweighted exploitation rates and ratios of large:small salmon were generated from the assessment rivers with retention angling fisheries. The unweighted averages were used as large rivers are now being dealt with independently. Population estimates for all rivers with counting facilities were included from their assessment information. In order to avoid double counting, the catches of rivers whose populations were included from assessments were subtracted from the total catch. In 1999, most of the Class IV rivers were in Bay St. George area of SFA 13 and the entire area returns and spawners were estimated based on assessments for 8 rivers expanded to the total drainage based on their proportionate contribution. In 2000, the rivers in Bay St. George were in three separate classes and were dealt with independently.

The mid-point of the estimated returns $(184,100)$ of 1 SW salmon to Newfoundland rivers in 2000 is $4 \%$ lower than 1999 and $8 \%$ higher than the average 1SW returns (170,300) for the period 1992-95 (Figure 4.2.2.1, Appendix 5). The 1992-95 1SW returns are higher than the returns in 1989-91, but similar to the returns to the rivers between 1971 and 1988. The mid-point $(9,300)$ of the estimated 2SW returns to Newfoundland rivers in 2000 remained the same as in 1999 (Figure 4.2.2.2, Appendix 5).

Québec: The mid-point $(30,800)$ of the estimated returns of 1 SW salmon to Québec in 2000 remained the same as the returns observed in 1999 and about the same as the 1995-99 average of 30,759 (Figure 4.2.2.1, Appendix 5).

The mid-point $(29,600)$ of the estimated returns of 2 SW salmon in Québec in 2000 is about the same as the returns observed for 1999 and a $21 \%$ decrease from the average of the years 1995-99 of 37,566 (Figure 4.2.2.2). Within the 1971-2000 time series, the 2000 value is the third lowest estimated and continues to decline from the high of 98,000 2SW salmon in 1980.

Gulf of St. Lawrence, SFAs 15-18: The mid-point $(51,900)$ of the estimated returns in 2000 of 1SW salmon returning to the Gulf of St. Lawrence was a $27 \%$ increase from 1999 and it is the third lowest value since 1984. The low values noted in 1997 through 2000 are low relative to the high value of about 189,000 in 1992 (Figure 4.2.2.1, Appendix 5).

The mid-point $(22,000)$ of the estimate of 2SW returns in 2000 is $5 \%$ higher than the estimate for 1999 and the fourth lowest of the time series (Figure 4.2.2.1, Appendix 5), the lowest being 1979 at 11,500. Returns of 2SW salmon have declined since 1995 with only slight improvement shown in 1999 and 2000.

Scotia-Fundy, SFAs 19-23: The mid-point $(14,700)$ of the estimate of the 1 SW returns in 2000 to the Scotia-Fundy Region was a $7 \%$ increase from the 1999 estimate, and the fifth lowest value in the time series, 1971-2000. Returns have generally been low since 1990 (Figure 4.2.2.1, Appendix 5).

The mid-point $(3,600)$ of the 2 SW returns in 2000 is $52 \%$ lower than the returns in 1999 and the lowest value in the time series, 1971-2000 (Figure 4.2.2.2, Appendix 5). A declining trend in returns has been observed from 1985 to 2000.

USA: Total salmon returns and spawners for USA rivers in 2000 were based on trap and weir catches (documented returns. Although some Maine rivers do not have fish counting facilities, a large portion of the total adult returns are documented and a method of estimating spawning escapement (and variance) in smaller Maine rivers based on redd counts is being developed. However, for 2000 the number of USA spawners was considered to be the same as the returns. The 1SW returns and spawners to USA rivers in 2000 were 270 fish. This was below the 1999 estimate and the previous 5 -year and 10-year averages. The 2SW returns and spawners to USA rivers in 2000 were 515 fish. This was also below the 1999 estimate and below the 5 -year and 10 -year averages. There were also 183 SW and repeat spawners in the USA returns.

### 4.2.3 Pre-fishery abundance estimates of non-maturing and maturing 1SW North American salmon

### 4.2.3.1 North American run-reconstruction model

The Working Group has used the North American run-reconstruction model to estimate pre-fishery abundance, which serves as the basis of abundance forecasts used in the provision of catch advice. The catch statistics used to derive returns and spawner estimates have been updated from those used in ICES 2000/ACFM:13 (Table 4.2.3.1). The North American run-reconstruction model has also been used to estimate the fishery exploitation rates for West Greenland and in homewaters.

### 4.2.3.2 Non-maturing 1SW salmon

The non-maturing component of 1 SW fish, destined to be 2 SW returns (excludes 3 SW and previous spawners) is represented by the pre-fishery abundance estimator for year i designated as [NN1(i)]. Definitions of the variables are given in Table 4.2.3.2. It is constructed by summing 2SW returns in year $\mathrm{i}+1$ [NR2(i+1)], 2SW salmon catches in commercial and native peoples' food fisheries in Canada [ $\mathrm{NC} 2(\mathrm{i}+1)$ ] and catches in year i from fisheries on nonmaturing 1SW salmon in Canada [ $\mathrm{NC1} 1$ (i)] and Greenland [NG1(i)]. In Labrador, native peoples' food harvests of small (AH_s) and large salmon (AH_l) were included in the reported catches for 1999-2000. Because harvests occurred in both Lake Melville and coastal areas of northern Labrador, the fraction of these catches that are immature was labeled as af_imm. This was necessary because non-maturing salmon do not occur in Lake Melville where approximately half the catch originated. However, non-maturing salmon may occur in coastal marine areas in the remainder of northern Labrador. Consequently, af_imm for the fraction of native peoples' harvests that were non-maturing was set at 0.05 to 0.1 which is half of $\mathrm{f}_{-} \mathrm{imm}$ from commercial fishery samples. The equations used to calculate NC 1 and NC 2 are as follows:

Eq. 4.2.3.1 $\mathrm{NCl}(\mathrm{i})=\left[\left(\mathrm{H}_{-} \mathrm{s}(\mathrm{i})_{\{1-7,14 b\}}+\mathrm{H}_{-} 1(\mathrm{i})_{\{1-7,14 b\}} * \mathrm{q}\right) * \mathrm{f}_{-} \mathrm{imm}\right]$

$$
+\left[\left(\mathrm{AH} \_\mathrm{s}(\mathrm{i})+\mathrm{AH} \_1(\mathrm{i}) * \mathrm{q}\right) * \text { af_imm }\right], \text { and }
$$

Eq. 4.2.3.2 $\mathrm{NC} 2(\mathrm{i}+1)=\left[\mathrm{H}_{-} 1(\mathrm{i}+1)_{\{1-7,14 \mathrm{~b}\}} *(1-\mathrm{q})\right]+\left[\mathrm{AH}_{-} 1(\mathrm{i}+1) *(1-\mathrm{q})\right]$
Similar to 1998 and 1999, the commercial fishery in Labrador remained closed in 2000. In past reports, salmon returns and spawners for Labrador which make up one of the six geographical areas contributing to NR2 for Canada were based on commercial fishery data. Since the commercial fishery was closed in Labrador in 1998, the time series also ended. However, in order to estimate pre-fishery abundance it was still necessary to include Labrador returns for 1998-2000. Consequently, a raising factor was developed by dividing pre-fishery abundance without Labrador into pre-fishery abundance with Labrador based on the time series of Labrador recruit estimates and pre-fishery abundance data from 1971-97. The raising factor (RFL2) to estimate returns to Labrador for 1998-2000 for 2SW salmon was set to the low and high range of values in the time series which was 1.05 to 1.27 . An assumed natural mortality rate [M] of 0.01 per month is used to adjust the numbers between the salmon fisheries on the 1 SW and 2 SW salmon ( 10 months) and between the fishery on 2 SW salmon and returns to the rivers ( 1 month) as shown below:

Eq. 4.2.3.3 $\mathrm{NN} 1(\mathrm{i})=[$ RFL2 $*((\mathrm{NR} 2(\mathrm{i}+1) / \mathrm{S} 1+\mathrm{NC} 2(\mathrm{i}+1)) / \mathrm{S} 2+\mathrm{NC} 1(\mathrm{i})]+\mathrm{NG} 1(\mathrm{i})$
where the parameters S 1 and S 2 are defined as $\exp (-\mathrm{M} * 1)$ and $\exp (-\mathrm{M} * 10)$, respectively. A detailed explanation of the model used to determine pre-fishery abundance is given in Rago et al. (1993a).

This estimated pre-fishery abundance represents the extant population and does not account for the fraction of the population present in a given fishery area. The model does not take into account non-catch fishing mortality in any of the fisheries. This is because rates for non-catch fishing mortality are not available on an annual basis and are not well described for some of the fisheries harvesting potential or actual 2SW salmon. Commercial catches were not included in the run-reconstruction model for the West Greenland fishery (1993 and 1994), Newfoundland fishery (1992-2000) and Labrador fishery (1998-2000) as these fisheries were closed.

As the pre-fishery abundance estimates for potential 2 SW salmon requires estimates of returns to rivers, the most recent year for which an estimate is available is 1999. This is because pre-fishery abundance estimates for 2000 require 2SW returns to rivers in North America in the year 2001 which of course are as of yet unavailable. The minimum and maximum values of the catches and returns for the 2 SW cohort are summarized in Table 4.2.3.3. The 1999 abundance estimates ranged between 57,800 and 130,436 salmon. The mid-point of this range $(94,118)$ is $2 \%$ higher than the 1998 value $(92,479)$ and is the $2^{\text {nd }}$ lowest in the 28 -year time series (Figure 4.2 .3 .1). The most recent three years are shown with hollow symbols as no Labrador values were estimated for these years and the raising factor described previously was used. The results indicate a slight levelling off of the general decline from 807,000 in 1975. The Working Group expressed concern about the continued decline in the pre-fishery abundance and its impact on spawner levels.

### 4.2.3.3 Maturing 1SW salmon

Estimation of an aggregate measure of abundance has utility for identifying trends, evaluating management measures, and investigating the influence of the marine environment on survival, distribution, and abundance of salmon. Maturing 1SW salmon are in some areas a major component of salmon stocks and measuring their abundance is thought to be important to provide measures of abundance of the entire cohort from a specific smolt class.

For the commercial catches in Newfoundland and Labrador, all small salmon are assumed to be 1SW fish based on catch samples which show the percentage of 1SW salmon to be in excess of $95 \%$. Large salmon are primarily MSW salmon but some maturing and non-maturing 1SW are also present in commercial catches in SFAs 1-7, 14B. Estimates of fractions of non-maturing salmon present in the Newfoundland and Labrador catch were presented in ICES 1991/Assess:12. The "large" category in SFAs 1-7 and 14B consists of 0.1-0.3 1SW salmon (Rago et al. 1993a; ICES 1993/Assess:10). Salmon catches in SFAs 8-14A are mainly maturing salmon (Idler et al. 1981). These values were assumed to apply to the Aboriginal food fishery catches in marine coastal areas of northern Labrador.

Similar to calculations to determine non-maturing 1SW salmon, a raising factor was also required to include Labrador returns in the maturing component of pre-fishery abundance necessitated by the closure of the commercial fishery in Labrador in 1998. Consequently, a raising factor was developed by dividing pre-fishery abundance without Labrador into pre-fishery abundance with Labrador based on the time series of Labrador recruit estimates and pre-fishery abundance data from 1971-97. The raising factor (RFL1) to estimate returns to Labrador for 1998-2000 for 1SW salmon was set to the low and high range of values in the time series which were 1.04 to 1.59 .

The maturing 1SW component is represented by the pre-fishery abundance estimator for year i [MN1(i)]. It is constructed by summing maturing 1SW returns in year i [MR1(i)] in Canada and the USA and catches in year i from commercial and food fisheries on maturing 1SW salmon in Newfoundland and Labrador [MC1(i)]. An assumed natural mortality rate [M] of 0.01 per month is used to adjust the numbers between the fishery on 1 SW salmon and returns to the rivers ( 1 month) as shown below:

Eq. 4.2.3.4 $\operatorname{MN1}(\mathrm{i})=[\mathrm{MR} 1(\mathrm{i}) / \mathrm{S} 1+\mathrm{MC1}(\mathrm{i})] *$ RFL1
where the parameter S 1 is defined as $\exp \left(-\mathrm{M}^{*} 1\right)$.

Eq. 4.2.3.5 $\mathrm{MCl}(\mathrm{i})=\left[\left(1-\mathrm{f} \_\mathrm{imm}\right)\left(\mathrm{H}_{-} \mathrm{s}(\mathrm{i})_{\{1-7,14 b\}}+\mathrm{q}^{*} \mathrm{H}_{-} \mathrm{l}(\mathrm{i})_{\{1-7,14 b\}}\right)\right]+\mathrm{H}_{-} \mathrm{s}(\mathrm{i})_{\{8-14 a\}}$

$$
+\left[\left(1-\mathrm{af} \_i m m\right)\left(A H \_s(i)+q^{*} A H \_1(i)\right)\right]
$$

This estimated pre-fishery abundance represents the extant population and does not account for the fraction of the population present in a given fishery area. The model does not take into account non-catch fishing mortality in any of the fisheries. This is because rates for non-catch fishing mortality are not available on an annual basis and are not well described for the fisheries harvesting 1SW salmon. Thus, catches used in the run-reconstruction model for the Newfoundland commercial fishery were set to zero for 1992-2000 and for Labrador for 1998-2000 to remain consistent with catches used in other years in these areas (see Section 4.1.1).

The minimum and maximum values of the catches and returns for the 1SW cohort are summarized in Table 4.2.3.4 and the mid-point values are shown in Figure 4.2.3.1. The most recent three years are shown with hollow symbols as no Labrador values were estimated for these years and the raising factor described previously was used. The mid-point of the range of pre-fishery abundance estimates for $2000(404,724)$ is $5 \%$ higher than in $1999(385,949)$ which had increased considerably from the low 1997 value of 325,433 which was the lowest estimated in the time series 19712000. The reduced values observed in 1978 and 1983-84 and 1994 were followed by large increases in pre-fishery abundance.

### 4.2.3.4 Total 1SW recruits (maturing and non-maturing)

Figure 4.2.3.1 shows the pre-fishery abundance of 1SW maturing and 1SW non-maturing salmon from North America for the period 1971 to 1999 and Figure 4.2.3.2 shows these data combined to give the total 1SW recruits. While maturing 1SW salmon in 1998-2000 have increased over the lowest value achieved in 1997, the non-maturing portion of these cohorts remained unchanged since 1997. Because the prefishery abundance has been consistently well below its conservation requirements, this situation is considered to be very serious. The decline in recruits in the time series is alarming. Although the declining trend appears common to both maturing and non-maturing portions of the cohort, nonmaturing 1SW salmon have declined further. The Working Group expressed concerns about these stock trends and recommended further investigation into their causes.

## 4.2 $4 \quad$ Spawning escapement and egg deposition

### 4.2.4.1 Egg depositions in rivers

On rivers not under colonization or rehabilitation, egg depositions in 2000 exceeded or equaled the river specific conservation requirements in 23 of the 54 assessed rivers ( $43 \%$ ) and were less than $50 \%$ of conservation in 18 other rivers ( $33 \%$ ) (Figure 4.2.4.1). Large deficiencies in egg depositions were noted in the Bay of Fundy and Atlantic coast of Nova Scotia where 10 of the 11 rivers assessed ( $91 \%$ ) had egg depositions which were less than $50 \%$ of conservation requirements. Proportionally fewer rivers in Gulf (27\%) and Québec ( $0 \%$ ) had egg depositions less than $50 \%$ of conservation. Only $27 \%$ of the Gulf rivers and $69 \%$ of the Québec rivers had egg depositions that equaled or exceeded conservation (Figure 4.2.4.1). In Newfoundland, $57 \%$ of the rivers assessed met or exceeded the conservation egg requirements and half of the others ( $21 \%$ ) had egg depositions that were less than $50 \%$ of requirement. The deficits occurred in the east and southwest rivers of Newfoundland (SFA 13) and in Labrador (Figure 4.2.4.1).

Fourteen rivers in Newfoundland (3) and Québec (11) are under rehabilitation or colonization programs where in recent years salmon have gained access to previously inaccessible habitat or to re-establish the wild production (Figure 4.2.4.1). Five of these rivers in Québec met or exceeded $90 \%$ of the conservation requirements in 2000. Egg depositions in $43 \%$ of these rivers were less than $50 \%$ of requirements. One international border river and four rivers in the USA are under rehabilitation. All of them had egg depositions less than $5 \%$ of conservation requirements.

Escapements over time relative to conservation requirements have improved in Newfoundland and decreased in all other areas of eastern Canada (Figure 4.2.4.2). The status of three Bay of Fundy/Atlantic coast of Nova Scotia rivers has severely declined, especially since 1991. The proportion of the conservation requirements achieved in 2000 was the lowest of the time series. For the Québec rivers, spawning escapements declined continually from a peak median value in 1988 with a slight recovery in 1995 and a similar increase in 1999. In almost all years inQuébec, the median proportion of conservation requirements achieved has exceeded the requirements. However, in 2000, the median proportion was the lowest value of the time series with $98 \%$ of the conservation requirement. The eight rivers of the Gulf of St. Lawrence have also been quite consistent in equalling or exceeding the conservation requirements but the median escapements were below conservation requirements in three of the last five years. Again in 2000, it was the lowest value ( $75 \%$ ) of the last fifteen years. Newfoundland rivers have shown the greatest improvement in the proportion of the spawning requirement achieved as a direct result of the commercial salmon and groundfish moratoria initiated in 1992. There was a decline in 1997 relative to 1996 but escapements increased again in 1998 (highest median values since the 1992), decreased only slightly in 1999 and were unchanged in 2000.

### 4.2.4.2 Run-reconstruction estimates of spawning escapement

Updated estimates for 2SW spawners were derived for the six geographic regions referenced in Section 4.2.2 (Table 4.2.4.1). Estimates of 1 SW spawners, 1971-2000 are provided in Table 4.2.4.2. These estimates were derived by subtracting the in-river removals from the estimates of returns to rivers. A comparison between the numbers of spawners, returns and spawning requirements for 1 SW and 2 SW salmon are shown in Figures 4.2.2.1 and 4.2.2.2 respectively (there are no spawning requirements defined specifically for 1 SW salmon).

Labrador: As previously explained, it was not possible to estimate spawners in Labrador in 1998-2000 due to lack of assessment information.

Newfoundland: The mid-point of the estimated numbers of 2 SW spawners $(9,000)$ in 2000 is about the same as that estimated in $1999(9,100)$ and is $224 \%$ of the total 2 SW spawner requirements for all rivers. The 2 SW spawner requirement has been met or exceeded in seven years since 1984 (Figure 4.2.2.2). The 1SW spawners $(167,200)$ in 2000 was about the same as the 169,700 in 1999 to 206,900 in 1999. The 1992-93, 1995-96 and 1998-2000 1SW spawners are higher than the spawners in 1989-91 and similar to levels in the late 1970s and 1980s (Figure 4.2.2.1). The spawning level in 1997 however was the third lowest in the data series, with 1989 and 1991 being lower. There had been a general increase in both 2SW and 1SW spawners during the period 1992-96 and 1998-2000 and this is consistent with the closure of the commercial fisheries in Newfoundland. For 1997, decreases occurred most strongly in the 1 SW spawners.

Québec: The mid-point of the estimated numbers of 2SW spawners $(19,700)$ in 2000 is $4 \%$ lower than that estimated in 1999 and is about $67 \%$ of the total 2 SW spawner requirements for all rivers (Figure 4.2.2.2). The spawning escapement in 2000 is the seventh lowest in the time series (1971-2000). Estimates of the numbers of spawners approximated the spawner requirement from 1971 to 1990 however they have been below requirements since 1990. The mid-point of the estimated 1SW spawners in $2000(22,100)$ was about $7 \%$ lower than 1999 (Figure 4.2.2.1). Spawning escapement of 1SW fish has generally been higher since the early 1980s than it was before this period.

Gulf of St. Lawrence: The mid-point of the estimated numbers of 2SW spawners $(19,800)$ in 2000 is about the same as that estimated in $1999(19,400)$ and is about $65 \%$ of the total 2 SW spawner requirements for all rivers in this region (Figure 4.2.2.2). This is the fifth time in ten years that these rivers have not exceeded their 2SW spawner requirements. The mid-point of the estimated spawning escapement of 1SW salmon $(34,800)$ increased by $28 \%$ from 1999 and is the ninth lowest in the time series, 1971-2000. The abundance remains low relative to the peak observed in 1992 (Figure 4.2.2.1). Spawning escapement has on average been higher in the mid-1980s than it was before and after this period.

Scotia-Fundy: The mid-point of the estimated numbers of 2SW spawners $(3,400)$ in 2000 is a $31 \%$ decrease from 1999 and is about $14 \%$ of the total 2 SW spawner requirements for rivers in this region (Figure 4.2.2.2). Neither the spawner estimates nor the spawner requirements include rivers of the inner Bay of Fundy (SFA 22 and part of SFA 23) as these rivers do not contribute to distant water fisheries and spawning escapements are extremely low. The 2SW spawning escapement in the rest of the area has been generally declining since 1985 and the 3,400 in 2000 is the lowest in the time series, 1971-2000. The mid-point of the estimated 1SW spawners $(14,300)$ in 2000 is a $7 \%$ increase from 1999 and is the eighth lowest in the time series, 1971-2000. There has been a general downward trend in 1SW spawners since 1990 (Figure 4.2.2.1).

USA:All age classes of spawners (1SW, 2SW, 3SW, and repeat) in 2000 ( 803 salmon) represented less than $2 \%$ of the 2 SW spawner requirements for all USA rivers. Spawning escapement of 2 SW salmon by river expressed as the percentage of conservation requirement was: $5 \%$ in the Penobscot, $2 \%$ in the Merrimack, and less than $1 \%$ in the Connecticut, Paucatuck, and other Maine rivers.

### 4.2.4.3 Escapement variability in North America

The projected numbers of potential 2SW spawners that could have returned to North America in the absence of fisheries can be computed from estimates of the pre-fishery abundance taking into consideration the 11 months of natural mortality at $1 \%$ per month. These values, termed potential 2 SW recruits, along with total North American 2SW returns, spawners and requirements are shown in Figure 4.2.4.3 and indicate that the overall North American spawner requirement could have been met, in the absence of all fisheries prior to, but not since 1994. The difference between the potential 2 SW recruits and actual 2 SW returns reflect the extent to which mixed stock fisheries at West Greenland and in SFAs 1-14 have reduced the populations.

Similarly, the impact of the Greenland fishery can be considered by subtracting the non-maturing 1SW salmon (accounting for natural mortality) harvested there from the total potential 2SW recruits. These values, termed 2SW recruits to North America, are also shown in Figure 4.2.4.3. The difference between the 2SW recruits to North America and the 2SW returns reflects the impact of removals by the commercial fisheries of Newfoundland and Labrador when they were open and the Labrador food fisheries since reports began in 1998. The 2SW recruits to North America indicate that, even if there had not been a West Greenland commercial fishery, spawner requirements could not have been met since 1992. The difference between the actual 2 SW returns and the spawner numbers reflects in-river removals throughout North America and coastal removals in Québec, Gulf and Scotia Fundy regions.

Following on the technique outlined in previous reports (ICES 1994/Assess:16, ICES 1995/Assess:14), the spawners in each geographic area were allocated (weighted forward) to the year of the non-maturing 1 SW component in the Northwest Atlantic using the weighted smolt age proportions from each area (Table 4.2.4.3). The total spawners for a given recruitment year in each area is the sum of the lagged spawners. Because the smolt age distributions in North America range from one to six years and the time series of estimated 2SW spawners to North America begins in 1971, the first recruiting year for which the total spawning stock size can be estimated is 1979 (although a value for 1978 was obtained by leaving out the 6 -year old smolt contribution which represents $4 \%$ of the Labrador stock complex (Table 4.2.4.3)). Since the 1999 2SW spawners to North America (except for Labrador) are known, the spawning stock contributing to the pre-fishery abundance up to 2002 is known for North America and up to 2003 except for Labrador (Figure 4.2.4.4, Table 4.2.4.4).

Spawning escapement of 2SW salmon to several stock complexes has been below the spawner requirement (Labrador, Québec, Scotia-Fundy, USA) since at least the 1980s (Figure 4.2.4.4). In the last four years, lagged spawner abundance has been increasing in Labrador and Newfoundland but decreasing in all other areas.

The relative contributions of the stocks from geographic area to the total spawning escapement of 2SW salmon has varied over time (Figure 4.2.4.5). The reduced potential contribution of Scotia-Fundy stocks and the increased proportion of the spawning stock from the Gulf of St. Lawrence and recently Labrador rivers to future recruitment is most evident. Thus production of non-maturing 1SW salmon would not be expected to increase dramatically from most areas of North America even if the sea survival improves. Only the Newfoundland stock complex has received
spawning escapements which have exceeded the area requirements, all other complexes were below requirement and some declined further in 2000.

### 4.2.5 Survival indices

Counts of smolts and adult salmon returns enable the estimation of indices of natural survival at sea, particularly following the closure of most northwest Atlantic commercial salmon fisheries in 1992. These estimates are potentially influenced by annual variation in size, age and sex composition of smolts leaving freshwater and possibly, annual variation in sea-age at maturity. There is information from 18 rivers in North America with smolt counts and corresponding adult counts. Data available in 2000 were from 10 wild and five hatchery populations distributed between Newfoundland (SFAs 4, 9, 11, 13, and 14a), Québec (Q2 and Q7), Nova Scotia (SFAs 20 and 21), New Brunswick (SFA 23) and Maine (USA).

Plots of survival rates over time (Figures 4.2.5.1 to 4.2.5.4) provide some insight into the impact of changes in management measures and possible changes in marine survival of wild and hatchery 1 SW and 2 SW stocks. In general the plots suggest:

- data sets that predate the commercial closures in 1992 and 1984 exhibit values that exceed those derived since fisheries closures, i.e., survival of North America stocks to home waters has not increased as expected
- $\quad 1 \mathrm{SW}$ survival greatly exceeds that of 2 SW fish, and
- survival of wild stocks exceeds that of hatchery stocks

Survival indices for 8 of 15 stocks returning 1SW fish in 2000 exceeded indices for 1SW fish in 1999. Six indices for 1SW fish decreased from 1999. None of the survival indices for nine stocks returning 2SW fish in 2000 increased from values in 1999. There have been no significant increasing trends ( $\mathrm{p} \leq 0.05$ ) in survival indices of any of the stock components since commercial closures in 1992.

| Sea-age \&stock | Province/region | Number of stocks |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Relative to 1999 |  |  | 9-Year Trend |  |  |
|  |  | 仑 | $\Leftrightarrow$ | (1) | へ | $\Leftrightarrow$ | (1) |
| 1SW Wild | West \& North Nfld | 1 |  | 2 |  | 3 |  |
|  | South Nfld | 3 |  |  |  | 3 |  |
|  | Québec |  | 1 | 1 |  | 2 |  |
|  | NS/NB | 1 |  | 1 |  | 1 |  |
| Hatchery | Québec |  |  | 1 |  | 1 |  |
|  | NS | 2 |  |  |  | 1 | 1 |
|  | NB |  |  | 1 |  | 1 |  |
|  | Maine | 1 |  |  |  | 1 |  |
|  | Total | 8 | 1 | 6 |  | 13 | 1 |
| 2SW Wild | South Nfld |  | 1 |  |  | 1 |  |
|  | Québec |  | 1 | 1 |  | 1 | 1 |
|  | NS |  |  | 1 |  |  |  |
| Hatchery | Québec |  | 1 |  |  | 1 |  |
|  | NS |  | 1 | 1 |  | 1 | 1 |
|  | NB |  |  | 1 |  | 1 |  |
|  | Maine |  |  | 1 |  |  | 1 |
|  | Total |  | 4 | 5 |  | 5 | 3 |

Catch advice is based on estimates of returns and spawners in home rivers and harvests in commercial fisheries (see Sections 4.2.2, 4.2.3, and 4.2.4). Escaped-farmed salmon have been most frequently found close to the principal salmon farming area of Passamaquoddy and Cobscook bays of the Bay of Fundy, although a few other farm sites occur in Nova Scotia and Newfoundland. (Figure 4.1.3.2).

The principal salmon farming industry in Bay the Fundy has grown extensively since 1984 since the closure of local commercial salmon fisheries. Estimates of returns and spawners in this area are based on assessments of wild and hatchery fish at counting facilities where escapes are identified on the basis of external characteristics and scale analysis and excluded from both the assessment and from ascending the rivers. Counts of wild/hatchery salmon in all the principal impacted rivers, (Table 4.1.3.1) generally total less than 200 fish in any year since 1990. Misclassification of many of the hatchery fish would be of little consequence to catch advice at even a regional scale.

Catch advice is not provided for inner Bay of Fundy rivers where some escapes have been observed. The occasional escape noted in other rivers of Nova Scotia and Newfoundland allows the possibility that escapes could influence angler harvests used to derive returns in some Salmon Fishing Areas. However the numbers of these fish must be of minor consequence to assessments. The occurrence of escapes in the West Greenland catch, the North American proportion of which is included in the total of North American production, has been investigated by Hansen et. al. (1997) and found to be less than one percent.

### 4.2.7 Summary of status of stocks in the North American Commission Area

Estimates of pre-fishery abundance suggest a continuing decline of North American adult salmon over the last 10 years. The total population of 1SW and 2SW Atlantic salmon in the northwest Atlantic has oscillated around a generally declining trend since the 1970s, and the abundance recorded in 1993-2000 was the lowest in the time series (Figure 4.2.3.2). During 1993 to 2000, the total population of 1 SW and 2SW Atlantic salmon was about one-half million fish, $45 \%$ of the average abundance during 1972 to 1990 . The decline has been more severe for the 2 SW salmon component than for the small salmon (maturing as 1 SW salmon) age group.

In most regions the returns of 2 SW fish are at or near the lower end of the 30-year time series (1971-2000) except Newfoundland where they are at the second highest level but are a minor age group component of the stocks in this area. Returns of 1SW salmon were at the lower end of the time series in Gulf, Scotia-Fundy, and USA and at about at the mid-point in Québec and Newfoundland.

The rank of the estimated returns in 2000 in the 1971-2000 time series for six regions in North America is shown below:

| Region | Rank of 2000 <br> 2000 time series | returns in <br> $(\mathbf{1}=$ highest $)$ | 1971- Mid-point estimate of 2SW spawners as <br> proportion of <br> escapement requirement <br> (\%) |
| :--- | :---: | :---: | :---: |
| Labrador | Unknown | Unknown | unknown |
| Newfoundland | 12 | 2 | 224 |
| Québec | 13 | 28 | 67 |
| Gulf | 22 | 27 | 65 |
| Scotia-Fundy | 24 | 30 | 14 |
| USA | 19 | 30 | 2 |

Trends in abundance of small salmon and large salmon within the geographic areas show a general synchronicity among the rivers. Returns of large salmon were generally decreased or were about the same as 1999 and among the lowest observed since 1987 while small salmon returns improved marginally for stocks in some areas. For the rivers of Newfoundland, large salmon returns were among the highest in the last 12 years but large salmon returns in the other areas were among the lowest. Large salmon in Newfoundland are predominantly repeat-spawning 1SW salmon while in other areas of eastern Canada, 2 SW and 3SW salmon make up varying proportions of the returns.

On rivers not under colonization or rehabilitation, egg depositions in 2000 exceeded or equaled the river specific conservation requirements in 23 of the 54 assessed rivers ( $43 \%$ ) and were less than $50 \%$ of conservation in 18 other rivers (33\%) (Figure 4.2.4.1). Large deficiencies in egg depositions were noted in the Bay of Fundy and Atlantic coast of Nova Scotia where 10 of the 11 rivers assessed ( $91 \%$ ) had egg depositions which were less than $50 \%$ of conservation
requirements. In 2000, the overall spawning escapement requirements for 2 SW salmon were not met in any area except Newfoundland. The overall 2SW spawning escapement requirement for Canada could have been met or exceeded in only nine (1974-78, 1980-82 and 1986) of the past 28 years (considering the mid-points of the estimates) by reduction of terminal fisheries (Figures 4.2.2.2 and 4.2.4.3). In the remaining years, spawning requirements could not have been met even if all terminal harvests had been eliminated. It is only within the last few years that Québec and the Gulf areas have failed to achieve their overall 2 SW salmon spawning requirements.

Measures of marine survival rates over time indicate that survival of North America stocks to home waters has not increased as expected as a result of fisheries changes. There have been no significant increasing trends in survival indices of any of the stock components since commercial closures in 1992.

Substantive increases in spawning escapements in recent years in northeast coast Newfoundland rivers and high smolt and juvenile production in many rivers, in conjunction with suitable ocean climate indices were suggestive of the potential for improved adult salmon returns for 1998 through 2000. Colder oceanic conditions both nearshore and in the Labrador Sea in the early 1990s are thought to have contributed to lower survival of salmon stocks in eastern Canada during that period. It was expected that increased marine water temperatures in 1994 to 1998 would have favoured marine survival and subsequent adult salmon production. Low returns of 2SW salmon in 1998 were consistent with the low 1SW returns of 1997.

Based on 1SW returns in 2000, no significant improvements in most areas, and further declines in some areas, are expected for large salmon in 2001. An additional concern is the low abundance levels which currently describe many salmon stocks in rivers in eastern Canada, particularly in the Bay of Fundy and Atlantic coast of Nova Scotia. USA salmon stocks exhibit these same downward trends. Most salmon rivers in the USA are hatchery-dependent and remain at low levels compared to conservation requirements. Despite major changes in fisheries management, returns have continued to decline in these areas and many populations are currently threatened with extirpation.

Fish passage efficiency, both upstream and downstream, limit populations at hydropower facilities such as on the Saint John, St. Croix, and Penobscot rivers. Aquaculture has continued to increase. Salmon populations of the Southern Upland of Nova Scotia have fallen to critically low levels in acid-impacted areas. Populations of Maine's downeast rivers are listed under the US Endangered Species Act and are suspected of having been adversely affected by freshwater habitat conditions and agriculture.

### 4.3 Effects on US and Canadian Stocks and Fisheries of Quota Management and Closure after 1991 in Canadian Commercial Salmon Fisheries, with special emphasis on the Newfoundland stocks

The Working Group previously considered the impact of the closure of the Newfoundland commercial fishery in 1992 on the Newfoundland stocks (ICES 1997/Assess:10).

Dempson et al. (1997) developed an index of salmon returns to illustrate the impact of the commercial salmon fishery moratorium on Newfoundland stocks. It is based on the difference between the returns prior to the moratorium (198491) when there was a commercial fishery to those in the years since the commercial fishery closed (1992-97). By averaging among rivers with counting facilities this provides an estimate of commercial fishing mortality which can then be used to estimate what returns would have been if the commercial fishery had not closed. The method assumes that natural mortality during the commercial fishery years remained at the same levels on average after the commercial fishery was closed. Average commercial fishing exploitation rate was $44 \%$ on small salmon and $75 \%$ on large. These exploitation rates should be regarded as a minimum values because it is evident that the natural component of marine survival has declined in recent years.

For 2SW salmon, if the commercial fishery had remained open during this period then, on average, from 1,942 to 6,821 fewer 2SW fish would have spawned. For 1SW salmon, had the commercial fishery remained open then, on average, from 37,672 to 96,655 fewer 1SW salmon would have spawned. For 2 SW salmon, in the years since the moratorium, spawner requirements have never been achieved if one uses the minimum estimates or have always been achieved using the maximum estimate. If the commercial fishery had not closed, then 2 SW spawners would never have achieved spawning requirements even at maximum estimates.

Within Newfoundland, the commercial fishery closure has resulted in increased escapements of both small and large salmon to rivers, higher catches of large salmon (which were subsequently released) in the recreational fishery, and increased spawning escapements of both size groups. These increased spawning escapements have not however always resulted in increased smolt production. Some areas of Newfoundland, particularly the south coast, did not see increases in escapement as was expected from the closure of the commercial fishery.

There are no changes recommended in the 2SW spawner requirements from those recommended previously. Spawner requirements for 2SW salmon for Canada now total 123,349 and for the USA, 29,199 for a combined total of 152,548 (Table 4.4.1). The Working Group again recommends that these requirements be refined as additional information on sea-age composition of spawners becomes available and as further understanding of life history strategies is gained.

### 4.5 Catch Options or Alternative Management Advice and Assessment of Risks Relative to the Objective of Exceeding Stock Conservation Limits

## Overview

Catch options are only provided for the non-maturing 1 SW and maturing 2 SW components as the maturing 1 SW component is not fished outside of home waters, and in the absence of significant marine interceptory fisheries, is managed in homewaters by the producing nations.

Catch histories of salmon which could have been available to the Greenland fishery, 1972-99, are provided in Tables 4.5.1 and 4.5.2. and expressed as 2 SW salmon equivalents. The Newfoundland-Labrador commercial fisheries historically harvested both maturing and non-maturing 1SW salmon as well as 2 SW maturing salmon. The harvest in these fisheries of repeat spawners and older sea-ages was not considered in the run reconstructions. Harvests of 1SW non-maturing salmon in Newfoundland-Labrador commercial fisheries have been adjusted by natural mortalities of 1\% per month for 11 months and 2 SW harvests in these same fisheries have been adjusted by one month to express all harvests as 2SW equivalents in the year and time they would reach rivers of origin. Starting in 1998, the Labrador commercial fishery was closed. A Native Peoples' fishery occurred in 1998-2000 which may have harvested, to some degree, mixed stocks and catches for this fishery have been included in Tables 4.5 .1 and 4.5 .2 . As well, a resident's food fishery in Labrador is included for the first time in 2000. Mortalities (principally in fisheries) in mixed stock and terminal fisheries areas in Canada are summed with those of USA to estimate total 2SW equivalent mortalities in North America (Table 4.5.1). Mortalities within North America peaked at about 382,000 in 1976 and are now about 14,400 2 SW salmon equivalents. In the most recent two years estimated (that is those since the closure of the Labrador commercial fishery), those taken as non-maturing fish in Labrador comprise 2\%, or less, of the total in North America.

Of the North American fisheries on the cohort destined to be 2 SW salmon, $90 \%$ of the catch comes from terminal fisheries in the most recent year. This value has ranged from as low as $19 \%$ in 1973, 1976 and 1987 to values of 76$91 \%$ in 1996-2000 fisheries (Table 4.5.1). The percentage increased significantly with the reduction and closures of the Newfoundland and Labrador commercial fisheries, particularly since 1992.

Table 4.5 .2 shows the mortalities expressed as 2SW equivalents in Canada, USA and Greenland for 1972-2000. Harvests within the USA of the total within North America approached $0.6 \%$ on a few occasions in the time series and as recently as in 1990. As well as these harvests in the USA, USA-origin salmon were also harvested in Canada during the time period indicated. The percentage of the total 2SW equivalents that has been taken in North American waters has ranged from $43-100 \%$, with the most recent year estimated at $75 \%$. The two years when $100 \%$ of the mortality occurred in North America were the years when the Greenland commercial fishery did not operate.

It is possible to provide catch advice for the North American Commission area for two years. The revised forecast for 2001 for 2SW maturing fish is based on a new forecast of the 2000 pre-fishery abundance and accounting for fish which were already removed from the cohort by fisheries in Greenland and Labrador in 2000 as 1SW non-maturing fish. The second is a new estimate for 2002 based on the pre-fishery abundance forecast for 2001 from Section 5.6. A consequence of these annual revisions is that the catch options for 2 SW equivalents in North America may change compared to the options developed the year before.

### 4.5.1 Catch advice for 2001 fisheries on 2SW maturing salmon

A revised forecast of the pre-fishery abundance for 2000 is provided in Table 5.6.1.1. This value of 225,708 is higher than the value forecast last year at this time of 179,897 (See Section 5.2 for more detailed derivation of the models used, etc.). A pre-fishery abundance of 225,708 in 2000 can be expressed as 2 SW equivalents by considering natural mortality of $1 \%$ per month for 10 months (a factor of 0.904837 ), resulting in 204,229 2 SW salmon equivalents. There have already been harvests of this cohort as 1SW non-maturing salmon in 2000 for both the Labrador (421) and Greenland ( 5,041 ) fisheries (Tables 4.5 .1 and 4.5.2) for a total of $5,4622 \mathrm{SW}$ salmon equivalents already harvested, when the mortality factor is considered.

Table 4.5.1.1 uses the probability density projections for the revised pre-fishery abundance estimate of 225,708 (at $50 \%$ probability) and subtracts the spawning reserve $(170,286)$ and the harvests in Greenland and Labrador of 1SW nonmaturing fish in 2000, and converts the remainder to 2 SW salmon equivalents. The calculation is as follows:
$\left[\mathrm{PFA}_{\mathrm{i}}-\right.$ spawning reserve - harvest in Greenland and Labrador in 2000 of 1 SW non-maturing fish $] \mathrm{x}$ exp $-(0.01 * 10$ months)]
where $\quad \mathrm{PFA}_{\mathrm{i}}=$ values from $25-75 \%$

$$
\text { spawning reserve }=170,286
$$

From Table 4.5.1.1, there are some harvest possibilities at forecasted levels which would be considered risk-neutral or risk-averse, that is, at probability levels of $50 \%$ and below down to about 14,000 fish at the $40 \%$ probability level. Any probability levels below this would suggest no harvest. The numbers provided for catch options refer to the composite North American fisheries. As the biological objective is to have all rivers reaching their conservation requirements, river-by-river management is necessary. On individual rivers, where spawning requirements are being achieved, there are no biological reasons to restrict the harvest.

To validate the forecasts provided by the PFA model, the Working Group considered historic information available on the returns of small and large salmon. Appropriate data were available for 20 rivers in Québec, six rivers of the Gulf, and three rivers of the Scotia-Fundy geographic area for the period 1991-1992 to 1999-2000, during which most commercial salmon fisheries were closed. The relationships between small salmon in year i and large salmon in year i+1 were significant for Québec and Scotia-Fundy (Figure 4.5.1). Forecasts of large salmon from the low returns of small salmon in 2000 (arrow on plots) indicate that there are no expectations for significant increases in large salmon in 2001.

There is no significant relationship between small and large salmon returns in the 'Gulf' rivers. However, the increasing proximity of recent data to the origin (Figure 4.5.1) and the low numbers of small salmon returns in 2000 also suggest low expectations for important increases in large salmon returns to the 'Gulf' in 2001. The evidence for each of the three geographic areas suggests that the PFA forecast of approximately 225,000 non-maturing fish in Greenland in 2000, i.e., triple the PFA value for the previous year and double the average PFA values of the previous 5 years, is highly unlikely.

Regional assessments in some areas of eastern North America provide a more detailed consideration of expectations for 2001, taking into consideration the contribution of all sea ages of salmon to the spawning population. By area, these are:

Labrador: Salmon returns in the year 2001 will be from a higher number of spawners than in recent years but the lack of long-term monitoring facilities makes it difficult to describe stock status or provide current expectations.

Newfoundland: The number of spawners has been relatively high in recent years, however, smolt output from all monitored rivers (with the exception of Highlands and Northeast Brook (Trepassey) in SFA 13) has declined in each of the past three years. In the absence of any improvement in marine survival rates, returns of small salmon in 2001 could be lower.

Québec : Returns of large salmon are expected to be adequate for the attainment of conservation requirements in 43 of the 44 salmon rivers in northern part of Québec; one river will remain closed. On the 74 salmon rivers in southern part of Québec, nine rivers will remained closed to fishing. Returns of large salmon are expected to be insufficient for attainment of conservation requirement on 34 rivers and consequently, only the retention of small salmon will be permitted on those rivers. Six other rivers are also under consideration for grilse-only retention.

Gulf: In SFA 15, returns in 2001 should approximate conservation requirements as they have in the last 5 years. Current levels of harvest have not been limiting the attainment of stock conservation. In SFA 16, neither large salmon nor eggs from small and large fish are expected to meet the conservation requirements in the Northwest and Southwest Miramichi, the Miramichi River overall, or the Buctouche River. It is expected, however that conservation requirements will be met or exceeded on the Tabusintac River. In SFA 18, Northumberland Strait and Cape Breton rivers, the Margaree, River Philip, and Sutherlands rivers are expected to meet or exceed conservation requirements. Stocks of the East River (Pictou), West River (Antigonish) and Wallace rivers are not expected to meet conservation requirements.

Scotia-Fundy: In SFAs 20-23, both large and small salmon returns, including hatchery supplements, are not expected to be adequate to meet conservation requirements. Many stocks are extirpated or at risk of extirpation and most rivers are
closed to even hook-and-release fishing. Few stocks of eastern Cape Breton (SFA 19) are expected to meet conservation requirements but are anticipated to be subjected to hook-and-release fishing.

USA: Salmon returns (both large and small) in 2000 are not expected to be sufficient to meet conservation requirements in any river, including those receiving hatchery stocking.

### 4.5.2 Catch advice for 2002 fisheries on 2SW maturing salmon

Most catches (93\%) in North America now take place in rivers or in estuaries. The commercial fisheries are now closed and the remaining coastal food fisheries in Labrador are mainly located close to river mouths and likely harvest few salmon from other than local rivers. Fisheries are principally managed on a river-by-river basis and in areas where retention of large salmon is allowed, it is closely controlled.

Catch options which could be derived from the prefishery abundance forecast for 2001 (295,678 at the $50 \%$ probability level) would apply principally to North American fisheries in 2002 and hence the level of fisheries in 2001 need to be accounted for before providing these catch options. Assuming probability values between 25 and $75 \%$, accounting for mortality and the spawning requirement and considering an allocation of $60 \%$ of the surplus to North America, would yield catch options in 2 SW salmon equivalents of 77,000 to 138,000 fish. The numbers provided for catch options refer to the composite North American fisheries. As the biological objective is to have all rivers reaching their conservation requirements, river-by-river management will be necessary. On individual rivers, where spawning requirements are being achieved, there are no biological reasons to restrict the harvest.

### 4.6 Data Deficiencies and Research Needs in the North American Commission Area

Some progress was made on research needs identified last year. The Working Group reiterates many of last year's recommendations and suggests some further ones. Relevant sections of this year's report are identified in parentheses.

1. Estimates of total returns to Labrador no longer exist. There is a critical need to develop alternate methods to derive estimates of salmon returns and develop habitat-based spawner requirements in Labrador, and to monitor salmon returns in the Ungava regions of Québec. (4.2.2; 4.2.4)
2. There is a need to investigate changes in the biological characteristics (mean weight, sex ratio, sea-age composition) of returns to rivers, spawning stocks of Canadian and US rivers, and the harvest in food fisheries in Labrador. These data and new information on measures of habitat and stock recruitment are necessary to re-evaluate existing estimates of spawner requirements in Canada and USA and for use in the run reconstruction model. (4.2.2; 4.2.3; 4.4)
3. There is a requirement for additional smolt-to-adult survival rates for wild salmon. As well, sea survival rates of wild salmon from rivers stocked with hatchery smolts should be examined to determine if hatchery return rates can be used as an index of sea survival of wild salmon elsewhere. (4.2.5)
4. Further basic research is needed on the spatial and temporal distribution of salmon and their predators at sea to assist in explaining variability in survival rates. (4.2.3; 4.2.5)
5. Return estimates for the few rivers (Annapolis, Cornwallis and Gaspareau) in SFA 22 that do contribute to distant fisheries should be developed and, when these are available, the SFA 22 spawning requirements for these rivers (476 fish) be included in the total.(4.4)
6. A consistent approach to estimating returns is needed, to incorporate broodstock, if offspring from such broodstock are stocked back into the management area from which their parents originated. (4.1.3)

Accounting of escaped-farmed salmon from North America indicates a high but undocumented mortality. Scale analysis of salmon captured at West Greenland indicated an infrequent appearance of escaped-farmed salmon. In order to substantiate this conclusion farmed-salmon need to be included in background genetic analysis and the data reexamined for the presence of escaped-farmed salmon of North American origin.

Table 4.1.2.1. Percentages by user group and province of small and large salmon harvested (by number) in the Atlantic salmon fisheries of eastern Canada during 2000.

|  |  | \% of Provincial Harvest |  |  | $\%$ of <br> eastern <br> Canada | Number of fish |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Native peoples' food fisheries | Recreational fisheries | Resident food fisheries |  |  |
| Small salmon |  |  |  |  |  |  |
| Newfoundland <br> Labrador | 1 | 17.5 | 74.4 | 8.1 | 45.5 | 22,813 |
| Québec |  | 15.3 | 84.7 | 0.0 | 13.1 | 6,588 |
| New Brunswick |  | 16.6 | 83.4 | 0.0 | 40.2 | 20,135 |
| P.E.I. |  | 15.6 | 84.4 | 0.0 | 0.4 | 179 |
| Nova Scotia |  | 27.0 | 73.0 | 0.0 | 0.8 | 393 |
| Large salmon |  |  |  |  |  |  |
| Newfoundland Labrador | / | 53.7 | 21.3 | 25.0 | 17.1 | 1,961 |
| Québec |  | 48.9 | 51.1 | 0.0 | 71.9 | 8,241 |
| New Brunswick |  | 100.0 | 0.0 | 0.0 | 9.1 | 1,047 |
| P.E.I. |  | - | - | - | 0.0 | 0 |
| Nova Scotia |  | 100.0 | 0.0 | 0.0 | 1.8 | 209 |
| Eastern Canada |  |  | \% by User Gro |  |  |  |
| Small salmon |  | 16.9 | 79.4 | 3.7 |  | 50,108 |
| Large salmon |  | 55.3 | 40.4 | 4.3 |  | 11,458 |

Table 4.1.2.2. Hook-and-released Atlantic salmon caught by recreational fishermen in Canada, 1984 - 2000.

| Year | Newfoundland |  |  | Nova Scotia |  |  | New Brunswick |  |  |  |  | Prince Edward Island |  |  | Quebec |  |  | CANADA* |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Small | Large | Total | Small | Large | Total | Small Kelt | Small Bright | $\begin{gathered} \text { Large } \\ \text { Kelt } \\ \hline \end{gathered}$ | Large Bright | Total | Small | Large | Total | Small | Large | Total | SMALL | LARGE | TOTAL |
| 1984 |  |  |  | 939 | 1,655 | 2,594 | 661 | 851 | 1,020 | 14,479 | 17,011 |  |  |  |  |  |  | 2,451 | 17,154 | 19,605 |
| 1985 |  | 315 | 315 | 1,323 | 6,346 | 7,669 | 1,098 | 3,963 | 3,809 | 17,815 | 26,685 |  |  | 67 |  |  |  | 6,384 | 28,285 | 34,669 |
| 1986 |  | 798 | 798 | 1,463 | 10,750 | 12,213 | 5,217 | 9,333 | 6,941 | 25,316 | 46,807 |  |  |  |  |  |  | 16,013 | 43,805 | 59,818 |
| 1987 |  | 410 | 410 | 1,311 | 6,339 | 7,650 | 7,269 | 10,597 | 5,723 | 20,295 | 43,884 |  |  |  |  |  |  | 19,177 | 32,767 | 51,944 |
| 1988 |  | 600 | 600 | 1,146 | 6,795 | 7,941 | 6,703 | 10,503 | 7,182 | 19,442 | 43,830 | 767 | 256 | 1,023 |  |  |  | 19,119 | 34,275 | 53,394 |
| 1989 |  | 183 | 183 | 1,562 | 6,960 | 8,522 | 9,566 | 8,518 | 7,756 | 22,127 | 47,967 |  |  |  |  |  |  | 19,646 | 37,026 | 56,672 |
| 1990 |  | 503 | 503 | 1,782 | 5,504 | 7,286 | 4,435 | 7,346 | 6,067 | 16,231 | 34,079 |  |  | 1,066 |  |  |  | 13,563 | 28,305 | 41,868 |
| 1991 |  | 336 | 336 | 908 | 5,482 | 6,390 | 3,161 | 3,501 | 3,169 | 10,650 | 20,481 | 1,103 | 187 | 1,290 |  |  |  | 8,673 | 19,824 | 28,497 |
| 1992 | 5,893 | 1,423 | 7,316 | 737 | 5,093 | 5,830 | 2,966 | 8,349 | 5,681 | 16,308 | 33,304 |  |  | 1,250 |  |  |  | 17,945 | 28,505 | 46,450 |
| 1993 | 18,196 | 1,731 | 19,927 | 1,076 | 3,998 | 5,074 | 4,422 | 7,276 | 4,624 | 12,526 | 28,848 |  |  |  |  |  |  | 30,970 | 22,879 | 53,849 |
| 1994 | 11,105 | 2,343 | 13,448 | 796 | 2,894 | 3,690 | 4,153 | 7,443 | 4,790 | 11,556 | 27,942 | 577 | 147 | 724 |  |  |  | 24,074 | 21,730 | 45,804 |
| 1995 | 12,383 | 2,588 | 14,971 | 979 | 2,861 | 3,840 | 770 | 4,260 | 880 | 5,220 | 11,130 | 209 | 139 | 348 |  | 922 | 922 | 18,601 | 12,610 | 31,211 |
| 1996 | 22,227 | 3,092 | 25,319 | 3,526 | 5,661 | 9,187 |  |  |  |  |  | 472 | 238 | 710 |  | 1,718 | 1,718 | 26,225 | 10,709 | 36,934 |
| 1997 | 17,362 | 3,810 | 21,172 | 717 | 3,358 | 4,075 | 3,457 | 4,870 | 3,786 | 8,874 | 20,987 | 210 | 118 | 328 | 182 | 1,643 | 1,825 | 26,798 | 21,589 | 48,387 |
| 1998 | 25,314 | 4,351 | 29,665 | 687 | 2,520 | 3,207 | 3,154 | 5,760 | 3,452 | 8,298 | 20,664 | 233 | 114 | 347 | 297 | 2,680 | 2,977 | 35,445 | 21,415 | 56,860 |
| 1999 | 18,119 | 4,534 | 22,653 | 591 | 2,161 | 2,752 | 3,155 | 5,631 | 3,456 | 8,281 | 20,523 | 192 | 157 | 349 | 298 | 2,693 | 2,991 | 27,986 | 21,282 | 49,268 |
| 2000 | 18,292 | 3,376 | 21,668 | 335 | 1,146 | 1,481 | 3,154 | 6,689 | 3,455 | 8,690 | 21,988 | 101 | 46 | 147 | 445 | 4,008 | 4,453 | 29,016 | 20,721 | 49,737 |

[^9]Table 4.1.3.1. Counts of salmon and percentage of the counts which were identified as aquaculture escapes (\% Aqua') at the counting facilities of the Magaguadavic River (SFA 23, Canada) and in rivers of eastern Maine, USA.

| Magaguadavic River (SFA 23, Canada) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1SW | \% Aqua' | MSW | \% Aqua | Total | \% Aqua’ |
| 1983 | 303 | - | 637 | - | 940 | - |
| 1984 | 249 | - | 534 | - | 783 | - |
| 1985 | 169 | - | 466 | - | 635 | - |
| 1988 | 291 | - | 398 | - | 689 | - |
| 1992 | 238 | 35 | 201 | 31 | 439 | 33 |
| 1993 | 208 | 46 | 177 | 29 | 385 | 38 |
| 1994 | 1064 | 94 | 228 | 73 | 1292 | 90 |
| 1995 | 540 | 90 | 198 | 85 | 738 | 89 |
| 1996 | 195 | 89 | 68 | 29 | 263 | 74 |
| 1997 | 94 | 63 | 47 | 49 | 141 | 58 |
| 1998 | 247 | 89 | 6 | 50 | 253 | 88 |
| 1999 | 74 | 74 | 29 | 83 | 103 | 77 |
| 2000 | 41 | 68 | 3 | 67 | 44 | 68 |


| Rivers of eastern Maine |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Union |  | St. Croix |  | Dennys |  | Narraguagus |  |
|  | Total Run | \% Aqua’ | Total run | \% Aqua' | Total run | $\%$ Aqua' | Total run | \% Aqua’ |
| 1994 | - | - | 181 | 54 | 47 | 89 | 52 | 2 |
| $1995^{1}$ | - | - | 60 | 22 | 9 | 44 | 56 | 0 |
| 1996 | - | - | 152 | 13 | 31 | 68 | 64 | 22 |
| 1997 | - | - | 70 | 39 | $2^{2}$ | 100 | 37 | 0 |
| 1998 | - | - | 65 | 37 | $1^{2}$ | 100 | 22 | 0 |
| 1999 | 72 | 91 | 36 | 64 | - | Unk | 35 | 8 |
| 2000 | 5 | 40 | 50 | 60 | 30 | 97 | 23 | 0 |

${ }^{1}$ High flows in 1995 may have affected accuracy of counts in all three rivers, especially the Dennys River
${ }^{2}$ Incomplete count of total run

Table 4.2.1.1. Comparison of returns of small salmon, large salmon, and size groups combined to assessed rivers of eastern Canada in 2000 relative to returns in 1999 and to returns in 1990 to 2000.

| Size group | Number of rivers in each category |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Returns in 2000 relative to returns in 1999 |  |  |  |
|  | Total | < $50 \%$ | $50 \%$ to $90 \%$ | $>=90 \%$ |
| Bay of Fundy and Atlantic Coast of Nova Scotia (SFA 19 to 23) |  |  |  |  |
| Small | 9 | 0 | 5 | 4 |
| Large | 9 | 4 | 2 | 3 |
| Small \& Large | 9 | 0 | 6 | 3 |
| Southern Gulf of St. Lawrence (SFA 15 to 18) |  |  |  |  |
| Small | 10 | 5 | 1 | 4 |
| Large | 10 | 3 | 3 | 4 |
| Small \& Large | 10 | 5 | 1 | 4 |
| Quebec (Zones Q1 to Q11) |  |  |  |  |
| Small | 13 | 0 | 3 | 10 |
| Large | 13 | 0 | 11 | 2 |
| Small \& Large | 28 | 1 | 14 | 13 |
| Newfoundland and Labrador (SFA 1 to 14) |  |  |  |  |
| Small | 21 | 2 | 8 | 11 |
| Large | 21 | 5 | 6 | 10 |
| Small \& Large | 21 | 1 | 7 | 13 |


| Size group | Number of rivers | Rank of 2000 within the 1990 to 2000 period (Rank 1 = highest) |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Best | Median | Worst |
| Bay of Fundy and Atlantic coast of Nova Scotia (SFA 19 to 23) |  |  |  |  |
| Small | 3 | 9 | 10 | 10 |
| Large | 4 | 9 | 11 | 11 |
| Small \& Large | 4 | 9 | 10 | 10 |
| Southern Gulf of St. Lawrence (SFA 15 to 18) |  |  |  |  |
| Small | 6 | 6 | 10.5 | 11 |
| Large | 6 | 6 | 10.5 | 11 |
| Small \& Large | 6 | 5 | 10.5 | 11 |
| Quebec (Zones Q1 to Q11) |  |  |  |  |
| Small | 11 | 2 | 6 | 11 |
| Large | 11 | 6 | 11 | 11 |
| Small \& Large | 25 | 1 | 9 | 11 |
| Newfoundland and Labrador (SFA 1 to 14) |  |  |  |  |
| Small | 8 | 2 | 6 | 9 |
| Large | 9 | 2 | 4 | 8 |
| Small \& Large | 8 | 2 | 6 | 8 |

Table 4.2.1.2. Index rivers in eastern North America with available juvenile abundance or smolt abundance estimates for 1971 to 2000. The index area refers to the SFAs or Zones which are assumed to be represented by the index rivers surveyed in those zones. River locations are shown in Figure 4.2.1.3.

| Geographic Area | SFA, Zons | Index river | AbundanceTyPe | Egg requirement (milions) |  | Index river! all inder rivers | River relative to SFA, Zont | $\begin{aligned} & \text { Roveras } \\ & \text { \% of Total } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | SFA, Zene | Index river |  |  |  |
| Labrader | 1 |  |  |  |  |  |  |  |
|  | 2 |  |  |  |  |  |  |  |
|  | 14B |  |  |  |  |  |  |  |
| Nemfoundiand | 3 |  |  |  |  |  |  |  |
|  | 4 | Campbeition | Smolat | 150.6 | 29 | 0.9\% | 18\% | 0.3\% |
|  | 5 |  |  | 37.9 |  |  |  |  |
|  | 8 |  |  |  |  |  |  |  |
|  | 7 |  |  |  |  |  |  |  |
|  | 8 |  |  |  |  |  |  |  |
|  | 9 | NE Trupassey | Smolts | 16.2 | 0.1 | 00\% | 0 0\% | 00\% |
|  |  | Roeky | Smola |  | 3.4 | 10\% | 14.2\% | 0.3\% |
|  | 10 |  |  | 7.8 |  |  |  |  |
|  | 11 | Corne | Smolt | 41.1 | 7.1.7 | 2.3\% | 1.0\% | 0.7\% |
|  |  | Littie | Smols |  | 0.3 | 0.1\% | 0.8\% | 0.0\% |
|  | 12 |  |  |  |  |  |  |  |
|  | 13 | Hightands | Smolas | 76.4 | 1.5 | 0.4\% | 2.0\% | 0.1\% |
|  | 14A | WAB | Smols | 19.1 | 0.9 | 0.3\% | 48\% | 0.1\% |
| Qulf | 15 | Restigouche | Juveniles | 71.9 | 53.6 | 15.8\% | 74.5\% | 6.1\% |
|  |  | Nepisiguit | duvenies |  | 8.5 | 2.8\% | 132\% | 0.9\% |
|  | 16 | Mramichi | Juvenies | 143.5 | 131.0 | 36.65 | 913\% | 12.4\% |
|  |  | Buatouche | Juveniles |  | 1.6 | 0.6\% | 1.1\% | 0.2\% |
|  | 16 | Margare | duvenies | 23.1 | 6.7 | 20\% | 290\% | 0.6\% |
|  |  | R. Philip | Juveniles |  | 2.3 | 0.7\% | 10.0\% | 0.2\% |
|  |  | Wallace | Juveniles |  | 1.5 | 0.4\% | 6.4\% | 0.1\% |
|  |  | Esst Pict | duveniles: |  | 1.1. | 0.5\% | 76\% | 0.2\% |
|  |  | West | Juveniles |  | 0.8 | 0.2\% | 3.5\% | 0.1\% |
| Sesta-Fundy | 19 | Mide | Juvenies | 21.2 | 21 | 0.6\% | 98\% | 0.2\% |
|  |  | Baddeck | Juveniles |  | 2.0 | 0.6\% | 9.4\% | 0.2\% |
|  |  | Nerth | duvenies |  | 0.9 | 0.3\% | 4.0\% | 0.1\% |
|  |  | Grand | Juveniles |  | 1.1 | 0.3\% | 5.2\% | 0.1\% |
|  |  | Inhabitarts | duveniles |  | 1.4 | 0.4\% | 6.5\% | 0.1\% |
|  | 20 | St Mary | Juveriles | 55.2 | 96 | 2.f\% | 173\% | 0.5\% |
|  | 21 | LaHaw | Juvenites | 77.6 | 12.2 | 3.6\% | 15.7\% | 1.2\% |
|  | 22 |  |  | 21.2 |  |  |  |  |
|  | 23 | Saint John | Juvenies | \$0.6 | 32.3 | 95\% | 35.7\% | 3.1\% |
|  |  | Nashwask | Juveniles |  | 13.7 | 4.0\% | 15.1\% | 1.3\% |
|  |  | Kenneberasis | duvenies |  | 8.0 | 1.5\% | 5.8\% | 0.5\% |
|  |  | Hammond | Juveniles |  | 4.0 | 12\% | 4.4\% | $0.0 \%$ |
| Quetec | 01 |  |  | 24.8 |  |  |  |  |
|  | g2 | Saint-Jean | Smols | 11.1 | 1.9 | 0.6\% | 169\% | 0.2\% |
|  | Q3 |  |  | 11.3 |  |  |  |  |
|  | 04 |  |  |  |  |  |  |  |
|  | 95 |  |  | 6.1 |  |  |  |  |
|  | Q6 |  |  | 5.8 |  |  |  |  |
|  | $g T$ | de la Trinite | Smoks | 14.1 | 1.6 | 0.8\% | 11.2\% | 0.2\% |
|  |  | Molsie | Smoke |  | 20.4 | 6.0\% | 145.1\% | 13\% |
|  | Q8 |  |  | 80.8 |  |  |  |  |
|  | 09 |  |  | 26.9 |  |  |  | 0.0\% |
|  | 910 | Bededelie | Smolat | 7.5 | 0.1 | 0.0\% | 1.9\% | 0,0\% |
| $\underline{\mathrm{E}}$. | Masine |  | duveniles | 5.5 | 5.5 | 1.6\% |  | 0.5\% |
|  |  |  |  |  |  |  |  |  |
| North Amerisa | Subtotal |  |  | 1063.0 | 339.4 | 1000\% |  | 32\% $2 \%$ |

Table 4.2.1.3. Number of rivers and percent of total zonal area represented by the indexed rivers in 1971 to 2000.

| Year | Rivers <br> Monitored | River Area as \% of <br> Total Indexed Area |
| :---: | :---: | :---: |
| 1971 | 2 | $11.5 \%$ |
| 1972 | 3 | $16.3 \%$ |
| 1973 | 3 | $16.3 \%$ |
| 1974 | 3 | $16.3 \%$ |
| 1975 | 4 | $16.9 \%$ |
| 1976 | 4 | $16.9 \%$ |
| 1977 | 5 | $17.1 \%$ |
| 1978 | 8 | $17.4 \%$ |
| 1979 | 6 | $18.1 \%$ |
| 1980 | 6 | $17.7 \%$ |
| 1981 | 10 | $20.9 \%$ |
| 1982 | 11 | $21.1 \%$ |
| 1983 | 9 | $20.8 \%$ |
| 1984 | 11 | $21.1 \%$ |
| 1985 | 11 | $20.9 \%$ |
| 1986 | 12 | $20.9 \%$ |
| 1987 | 13 | $21.8 \%$ |
| 1988 | 12 | $21.2 \%$ |
| 1989 | 12 | $20.5 \%$ |
| 1990 | 16 | $22.7 \%$ |
| 1991 | 18 | $23.3 \%$ |
| 1992 | 19 | $23.5 \%$ |
| 1993 | 23 | $26.8 \%$ |
| 1994 | 24 | $26.8 \%$ |
| 1995 | 28 | $29.2 \%$ |
| 1996 | 26 | $28.1 \%$ |
| 1997 | 29 | $29.3 \%$ |
| 1998 | 30 | $29.7 \%$ |
| 1999 | 27 | $27.6 \%$ |
| 2000 | 26 | $26.6 \%$ |
|  |  |  |

Table 4.2.2.1 Estimated numbers of 1SW returns in North America by geographic regions, 1971 - 2000.

| Year | Labrador |  | Newfoundland |  | Quebec |  | Gulf of St. Lawrence |  | Scotia-Fundy |  | USA | North America |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max |  | Min | Max | Mid-points |
| 1971 | 32,966 | 115,382 | 112,644 | 226,129 | 14,969 | 22,453 | 33,118 | 57,973 | 11,515 | 19,525 | 32 | 205,245 | 441,495 | 323,370 |
| 1972 | 24,675 | 86,362 | 109,282 | 219,412 | 12,470 | 18,704 | 42,202 | 73,711 | 9,522 | 16,915 | 18 | 198,169 | 415,122 | 306,645 |
| 1973 | 5,399 | 18,897 | 144,267 | 289,447 | 16,585 | 24,877 | 43,681 | 77,102 | 14,766 | 24,823 | 23 | 224,721 | 435,169 | 329,945 |
| 1974 | 27,034 | 94,619 | 85,216 | 170,748 | 16,791 | 25,186 | 65,673 | 114,083 | 26,723 | 44,336 | 55 | 221,491 | 449,026 | 335,259 |
| 1975 | 53,660 | 187,809 | 112,272 | 225,165 | 18,071 | 27,106 | 58,613 | 101,887 | 25,940 | 36,316 | 84 | 268,639 | 578,367 | 423,503 |
| 1976 | 37,540 | 131,391 | 115,034 | 230,595 | 19,959 | 29,938 | 90,308 | 155,693 | 36,931 | 55,937 | 186 | 299,958 | 603,740 | 451,849 |
| 1977 | 33,409 | 116,931 | 110,114 | 220,501 | 18,190 | 27,285 | 31,322 | 56,088 | 30,860 | 48,387 | 75 | 223,971 | 469,268 | 346,619 |
| 1978 | 16,155 | 56,542 | 97,375 | 195,048 | 16,971 | 25,456 | 26,008 | 45,413 | 12,457 | 16,587 | 155 | 169,121 | 339,201 | 254,161 |
| 1979 | 21,943 | 76,800 | 107,402 | 215,160 | 21,683 | 32,524 | 50,872 | 93,340 | 30,875 | 49,052 | 250 | 233,025 | 467,126 | 350,075 |
| 1980 | 49,670 | 173,845 | 121,038 | 242,499 | 29,791 | 44,686 | 45,716 | 81,737 | 49,925 | 73,560 | 818 | 296,958 | 617,145 | 457,051 |
| 1981 | 55,046 | 192,662 | 157,425 | 315,347 | 41,667 | 62,501 | 70,238 | 128,658 | 37,371 | 62,083 | 1,130 | 362,877 | 762,381 | 562,629 |
| 1982 | 38,136 | 133,474 | 141,247 | 283,002 | 23,699 | 35,549 | 79,874 | 143,543 | 23,839 | 38,208 | 334 | 307,129 | 634,111 | 470,620 |
| 1983 | 23,732 | 83,061 | 109,934 | 220,216 | 17,987 | 26,981 | 25,337 | 43,922 | 15,553 | 23,775 | 295 | 192,838 | 398,250 | 295,544 |
| 1984 | 12,283 | 42,991 | 130,836 | 262,061 | 21,566 | 30,894 | 37,696 | 63,943 | 27,954 | 47,493 | 598 | 230,933 | 447,980 | 339,456 |
| 1985 | 22,732 | 79,563 | 121,731 | 243,727 | 22,771 | 33,262 | 61,255 | 110,580 | 29,410 | 51,983 | 392 | 258,290 | 519,507 | 388,899 |
| 1986 | 34,270 | 119,945 | 125,329 | 251,033 | 33,758 | 46,937 | 114,718 | 204,455 | 30,935 | 54,678 | 758 | 339,768 | 677,807 | 508,787 |
| 1987 | 42,938 | 150,283 | 128,578 | 257,473 | 37,816 | 54,034 | 86,564 | 156,086 | 31,746 | 55,564 | 1,128 | 328,770 | 674,567 | 501,668 |
| 1988 | 39,892 | 139,623 | 133,237 | 266,895 | 43,943 | 62,193 | 123,578 | 223,368 | 32,992 | 56,935 | 992 | 374,635 | 750,007 | 562,321 |
| 1989 | 27,113 | 94,896 | 60,260 | 120,661 | 34,568 | 48,407 | 72,944 | 129,515 | 34,957 | 59,662 | 1,258 | 231,101 | 454,400 | 342,750 |
| 1990 | 15,853 | 55,485 | 99,543 | 199,416 | 39,962 | 54,792 | 83,670 | 159,455 | 33,939 | 60,828 | 687 | 273,654 | 530,664 | 402,159 |
| 1991 | 12,849 | 44,970 | 64,552 | 129,308 | 31,488 | 42,755 | 59,721 | 113,722 | 19,759 | 31,555 | 310 | 188,679 | 362,619 | 275,649 |
| 1992 | 17,993 | 62,094 | 118,778 | 237,811 | 35,257 | 48,742 | 146,539 | 231,291 | 22,832 | 37,340 | 1,194 | 342,594 | 618,473 | 480,533 |
| 1993 | 25,186 | 80,938 | 134,150 | 268,550 | 30,645 | 42,156 | 89,934 | 146,977 | 16,714 | 27,539 | 466 | 297,095 | 566,627 | 431,861 |
| 1994 | 18,159 | 56,888 | 95,981 | 192,138 | 29,667 | 40,170 | 55,639 | 117,549 | 8,216 | 11,583 | 436 | 208,098 | 418,763 | 313,430 |
| 1995 | 25,022 | 76,453 | 202,739 | 435,153 | 23,851 | 32,368 | 26,019 | 96,871 | 14,239 | 21,822 | 213 | 292,082 | 662,880 | 477,481 |
| 1996 | 51,867 | 153,553 | 257,215 | 559,079 | 32,008 | 42,558 | 50,313 | 99,615 | 22,795 | 36,047 | 651 | 414,848 | 891,504 | 653,176 |
| 1997 | 66,812 | 155,963 | 99,029 | 146,050 | 24,300 | 33,018 | 27,515 | 54,511 | 7,173 | 10,467 | 365 | 225,194 | 400,374 | 312,784 |
| 1998 | - | - | 146,371 | 247,035 | 24,029 | 33,524 | 38,029 | 69,155 | 16,770 | 26,481 | 403 | - | - | - |
| 1999 | - | - | 156,740 | 224,959 | 25,639 | 36,296 | 28,867 | 53,244 | 10,556 | 16,901 | 419 | - | - | - |
| 2000 | - | - | 116,167 | 252,102 | 25,216 | 36,309 | 40,215 | 63,624 | 10,997 | 18,343 | 270 | - | - | - |

Labrador: SFAs $1,2 \& 14 B$
Newfoundland: SFAs 3-14A
Gulf of St. Lawrence: SFAs 15-18
Scotia-Fundy: SFAs 19-23 (SFA 22 is not included as it does not produce 2SW salmon)
Quebec: Q1-Q11

Table 4.2.2.2 Estimated numbers of 2SW returns in North America by geographic regions, 1971-2000.

|  | Labrador |  | Newfoundland |  | Quebec |  | Gulf of St. Lawrence |  | Scotia-Fundy |  | USA | North America |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max |  | Min | Max | Mid-points |
| 1971 | 4,312 | 29,279 | 2,388 | 8,923 | 34,568 | 51,852 | 29,483 | 46,831 | 11,187 | 16,410 | 653 | 81,937 | 153,295 | 117,616 |
| 1972 | 3,706 | 25,168 | 2,511 | 9,003 | 45,094 | 67,642 | 35,640 | 59,937 | 14,028 | 19,731 | 1,383 | 102,364 | 182,865 | 142,614 |
| 1973 | 5,183 | 35,196 | 2,995 | 11,527 | 49,765 | 74,647 | 34,911 | 59,550 | 10,359 | 14,793 | 1,427 | 104,641 | 197,140 | 150,890 |
| 1974 | 5,003 | 34,148 | 1,940 | 6,596 | 66,762 | 100,143 | 49,081 | 83,402 | 21,902 | 29,071 | 1,394 | 146,082 | 254,754 | 200,418 |
| 1975 | 4,772 | 32,392 | 2,305 | 7,725 | 56,695 | 85,042 | 31,175 | 51,864 | 23,944 | 31,496 | 2,331 | 121,222 | 210,851 | 166,036 |
| 1976 | 5,519 | 37,401 | 2,334 | 7,698 | 56,365 | 84,547 | 29,266 | 51,427 | 21,768 | 29,837 | 1,317 | 116,569 | 212,228 | 164,398 |
| 1977 | 4,867 | 33,051 | 1,845 | 6,247 | 66,442 | 99,663 | 58,822 | 100,766 | 28,606 | 39,215 | 1,998 | 162,581 | 280,941 | 221,761 |
| 1978 | 3,864 | 26,147 | 1,991 | 6,396 | 59,826 | 89,739 | 30,465 | 51,481 | 16,946 | 22,561 | 4,208 | 117,301 | 200,531 | 158,916 |
| 1979 | 2,231 | 15,058 | 1,088 | 3,644 | 32,994 | 49,491 | 8,671 | 14,324 | 8,962 | 12,968 | 1,942 | 55,888 | 97,427 | 76,658 |
| 1980 | 5,190 | 35,259 | 2,432 | 7,778 | 78,447 | 117,670 | 43,407 | 73,841 | 31,897 | 44,823 | 5,796 | 167,169 | 285,167 | 226,168 |
| 1981 | 4,734 | 32,051 | 3,451 | 12,035 | 61,633 | 92,449 | 17,743 | 29,594 | 19,030 | 28,169 | 5,601 | 112,192 | 199,900 | 156,046 |
| 1982 | 3,491 | 23,662 | 2,914 | 9,012 | 54,655 | 81,982 | 31,652 | 51,128 | 17,516 | 24,182 | 6,056 | 116,284 | 196,022 | 156,153 |
| 1983 | 2,538 | 17,181 | 2,586 | 8,225 | 44,886 | 67,329 | 29,038 | 46,874 | 14,310 | 20,753 | 2,155 | 95,513 | 162,517 | 129,015 |
| 1984 | 1,806 | 12,252 | 2,233 | 7,060 | 44,661 | 59,160 | 20,478 | 34,131 | 17,938 | 27,899 | 3,222 | 90,339 | 143,724 | 117,031 |
| 1985 | 1,448 | 9,779 | 958 | 3,059 | 45,916 | 61,460 | 23,106 | 43,533 | 22,841 | 38,784 | 5,529 | 99,798 | 162,144 | 130,971 |
| 1986 | 2,470 | 16,720 | 1,606 | 5,245 | 55,159 | 72,560 | 36,214 | 70,921 | 18,102 | 33,101 | 6,176 | 119,727 | 204,723 | 162,225 |
| 1987 | 3,289 | 22,341 | 1,336 | 4,433 | 52,699 | 68,365 | 22,668 | 47,919 | 11,529 | 20,679 | 3,081 | 94,602 | 166,818 | 130,710 |
| 1988 | 2,068 | 14,037 | 1,563 | 5,068 | 56,870 | 75,387 | 26,140 | 49,956 | 10,370 | 19,830 | 3,286 | 100,297 | 167,564 | 133,930 |
| 1989 | 2,018 | 13,653 | 697 | 2,299 | 51,656 | 67,066 | 17,311 | 35,338 | 11,939 | 21,818 | 3,197 | 86,819 | 143,371 | 115,095 |
| 1990 | 1,148 | 7,790 | 1,347 | 4,401 | 50,261 | 66,352 | 24,616 | 53,110 | 10,248 | 18,871 | 5,051 | 92,671 | 155,576 | 124,123 |
| 1991 | 548 | 3,740 | 1,054 | 3,429 | 46,841 | 60,724 | 20,983 | 44,446 | 10,613 | 17,884 | 2,647 | 82,687 | 132,871 | 107,779 |
| 1992 | 2,515 | 15,548 | 3,111 | 10,554 | 46,917 | 61,285 | 29,101 | 61,122 | 9,777 | 16,456 | 2,459 | 93,880 | 167,425 | 130,652 |
| 1993 | 3,858 | 18,234 | 1,499 | 5,094 | 37,023 | 46,484 | 25,753 | 51,793 | 6,764 | 11,087 | 2,231 | 77,128 | 134,924 | 106,026 |
| 1994 | 5,653 | 24,396 | 1,902 | 6,174 | 37,703 | 47,180 | 22,097 | 57,055 | 4,379 | 6,908 | 1,346 | 73,080 | 143,058 | 108,069 |
| 1995 | 12,368 | 44,205 | 3,635 | 12,592 | 43,755 | 54,186 | 24,276 | 62,950 | 4,985 | 8,317 | 1,748 | 90,767 | 183,998 | 137,382 |
| 1996 | 9,113 | 32,759 | 4,457 | 14,159 | 39,413 | 49,846 | 20,380 | 42,964 | 7,227 | 12,054 | 2,407 | 82,996 | 154,188 | 118,592 |
| 1997 | 9,384 | 23,833 | 3,887 | 8,355 | 32,443 | 41,017 | 17,563 | 37,804 | 3,645 | 5,922 | 1,611 | 68,533 | 118,542 | 93,537 |
| 1998 | - | - | 5,322 | 12,453 | 24,295 | 31,726 | 8,260 | 19,498 | 2,728 | 6,003 | 1,526 | - | - | - |
| 1999 | - | - | 4,254 | 14,262 | 25,362 | 33,622 | 12,389 | 29,399 | 3,482 | 7,107 | 1,168 | - | - | - |
| 2000 | - | - | 2,218 | 16,360 | 24,695 | 34,543 | 14,052 | 29,955 | 2,038 | 5,079 | 533 | - | - | - |

[^10]Table 4.2.3.1 Run reconstruction data inputs for harvests used to estimate pre-fishery abundance of maturing and nonmaturing 1SW salmon of North American origin (terms defined in Table 4.2.3.2).

| 1SW <br> Year <br> (i) | AH_Small (i) | $\left\lvert\, \begin{gathered} \{1\} \\ \mathrm{AH}_{-} \mathrm{Large} \\ (\mathrm{i}+1) \end{gathered}\right.$ | AH_Large (i) | $\{1-7,14 b$ <br> Small <br> (i) |  | $\begin{array}{\|c} \hline\{8-14 \mathrm{a}\} \\ \mathrm{H}_{-} \mathrm{Small} \\ \text { (i) } \end{array}$ | $\begin{array}{\|r} \text { H_Large } \\ (\mathrm{i}+1) \\ \hline \end{array}$ | $\begin{array}{\|c} \{1-7,14 \mathrm{~b}\} \\ \text { H_Large } \\ (\mathrm{i}+1) \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 0 | 0 | 0 | 158896 | 199176 | 70936 | 42861 | 144496 |
| 1972 | 0 | 0 | 0 | 143232 | 144496 | 111141 | 43627 | 227779 |
| 1973 | 0 | 0 | 0 | 188725 | 227779 | 176907 | 85714 | 196726 |
| 1974 | 0 | 0 | 0 | 192195 | 196726 | 153278 | 72814 | 215025 |
| 1975 | 0 | 0 | 0 | 302348 | 215025 | 91935 | 95714 | 210858 |
| 1976 | 0 | 0 | 0 | 221766 | 210858 | 118779 | 63449 | 231393 |
| 1977 | 0 | 0 | 0 | 220093 | 231393 | 57472 | 37653 | 155546 |
| 1978 | 0 | 0 | 0 | 102403 | 155546 | 38180 | 29122 | 82174 |
| 1979 | 0 | 0 | 0 | 186558 | 82174 | 62622 | 54307 | 211896 |
| 1980 | 0 | 0 | 0 | 290127 | 211896 | 94291 | 38663 | 211006 |
| 1981 | 0 | 0 | 0 | 288902 | 211006 | 60668 | 35055 | 129319 |
| 1982 | 0 | 0 | 0 | 222894 | 129319 | 77017 | 28215 | 108430 |
| 1983 | 0 | 0 | 0 | 166033 | 108430 | 55683 | 15135 | 87742 |
| 1984 | 0 | 0 | 0 | 123774 | 87742 | 52813 | 24383 | 70970 |
| 1985 | 0 | 0 | 0 | 178719 | 70970 | 79275 | 22036 | 107561 |
| 1986 | 0 | 0 | 0 | 222671 | 107561 | 91912 | 19241 | 146242 |
| 1987 | 0 | 0 | 0 | 281762 | 146242 | 82401 | 14763 | 86047 |
| 1988 | 0 | 0 | 0 | 198484 | 86047 | 74620 | 15577 | 85319 |
| 1989 | 0 | 0 | 0 | 172861 | 85319 | 60884 | 11639 | 59334 |
| 1990 | 0 | 0 | 0 | 104788 | 59334 | 46053 | 10259 | 39257 |
| 1991 | 0 | 0 | 0 | 89099 | 39257 | 42721 | 0 | 32341 |
| 1992 | 0 | 0 | 0 | 24249 | 32341 | 0 | 0 | 17096 |
| 1993 | 0 | 0 | 0 | 17074 | 17096 | 0 | 0 | 15377 |
| 1994 | 0 | 0 | 0 | 8640 | 15377 | 0 | 0 | 11176 |
| 1995 | 0 | 0 | 0 | 7980 | 11176 | 0 | 0 | 7272 |
| 1996 | 0 | 0 | 0 | 7849 | 7272 | 0 | 0 | 6943 |
| 1997 | 0 | 2269 | 0 | 9753 | 6943 | 0 | 0 | 0 |
| 1998 | 2988 | 1084 | 2269 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 2739 | 1545 | 1084 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 5842 | 0 | 1545 | 0 | 0 | 0 | 0 | 0 |

Table 4.2.3.2 Definitions of key variables used in continental run-reconstruction models for North American salmon.

## Variable Definition

| i | Year of the fishery on 1SW salmon in Greenland and Canada |
| :---: | :---: |
| M | Natural mortality rate (0.01 per month) |
| t1 | Time between the mid-point of the Canadian fishery and return to river $=2$ months |
| S1 | Survival of 1SW salmon between the homewater fishery and return to river $\{\exp (-\mathrm{Mt} 1)\}$ |
| H_s(i) | Number of "Small" salmon caught in Canada in year i; fish $<2.7 \mathrm{~kg}$ |
| H_l(i) | Number of "Large" salmon caught in Canada in year i; fish $>=2.7 \mathrm{~kg}$ |
| AH_s | Aboriginal and resident food harvests of small salmon in northern Labrador |
| AH_1 | Aboriginal and resident food harvest of large salmon in northern Labrador |
| f_imm | Fraction of 1SW salmon that are immature, i.e. non-maturing: range $=0.1$ to 0.2 |
| af_imm | Fraction of 1SW salmon that are immature in native and resident food fisheries in N Labrador |
| q | Fraction of 1SW salmon present in the large size market category; range $=0.1$ to 0.3 |
| MC1(i) | Harvest of maturing 1SW salmon in Newfoundland and Labrador in year i |
| i +1 | Year of fishery on 2SW salmon in Canada |
| MR1(i) | Return estimates of maturing 1SW salmon in Atlantic Canada in year i |
| NN1(i) | Pre-fishery abundance of non-maturing 1SW + maturing 2SW salmon in year i |
| NR(i) | Return estimates of non-maturing + maturing 2SW salmon in year i |
| NR2(i+1) | Return estimates of maturing 2SW salmon in Canada |
| $\mathrm{NC1}(\mathrm{i})$ | Harvest of non-maturing 1SW salmon in Nfld + Labrador in year i |
| NC2(i+1) | Harvest of maturing 2SW salmon in Canada |
| NG(i) | Catch of 1SW North American origin salmon at Greenland |
| S2 | Survival of 2SW salmon between Greenland and homewater fisheries |
| MN1(i) | Pre-fishery abundance of maturing 1SW salmon in year I |
| RFL1 | Labrador raising factor for 1SW used to adjust pre-fishery abundance |
| RFL2 | Labrador raising factor for 2SW used to adjust pre-fishery abundance |

Table 4.2.3.3 Run reconstruction data inputs used to estimate pre-fishery abundance of non-maturing (NN1) 1SW salmon of North American origin (terms defined in Table 4.2.3.2).

| 1SW <br> Year (i) | NG1 <br> (i) | $\begin{aligned} & \mathrm{NC1} \\ & \mathrm{~min} \end{aligned}$ <br> (i) | $\begin{aligned} & \max \\ & (\mathrm{i}) \\ & \hline \end{aligned}$ | $\begin{aligned} & \mathrm{NC} 2 \\ & \mathrm{~min} \\ & (\mathrm{i}+1) \end{aligned}$ | $\begin{aligned} & \max \\ & (\mathrm{i}+1) \\ & \hline \end{aligned}$ | $\begin{aligned} & \mathrm{NR} 2 \\ & \mathrm{~min} \\ & (\mathrm{i}+1) \end{aligned}$ | $\begin{aligned} & \max \\ & (\mathrm{i}+1) \\ & \hline \end{aligned}$ | $\begin{aligned} & \mathrm{NN} 1 \\ & \min \\ & \text { (i) } \\ & \hline \end{aligned}$ | $\begin{aligned} & \max \\ & (\mathrm{i}) \\ & \hline \end{aligned}$ | midpoint (i) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 17881 | 43730 | 14008 | 172907 | 102364 | 182865 |  |  | -652798 |
| 1971 | 287672 | 17881 | 43730 | 144008 | 17290 | 102364 | 182865 | 578974 | 72662 | 652798 |
| 1972 | 200784 | 15768 | 37316 | 203072 | 248628 | 104641 | 197140 | 557790 | 732940 | 645365 |
| 1973 | 241493 | 21150 | 51412 | 223422 | 262767 | 146082 | 254754 | 672631 | 867684 | 770157 |
| 1974 | 220584 | 21187 | 50243 | 223332 | 266337 | 121222 | 210851 | 623907 | 800542 | 712224 |
| 1975 | 278839 | 32385 | 73371 | 243315 | 285486 | 116569 | 212228 | 710252 | 904626 | 807439 |
| 1976 | 155896 | 24285 | 57005 | 225424 | 271703 | 162581 | 280941 | 610799 | 826787 | 718793 |
| 1977 | 189709 | 24323 | 57902 | 146535 | 177644 | 117301 | 200531 | 506919 | 667787 | 587353 |
| 1978 | 118853 | 11796 | 29813 | 86644 | 103079 | 55888 | 97427 | 288792 | 371342 | 330067 |
| 1979 | 200061 | 19478 | 42242 | 202634 | 245013 | 167169 | 285167 | 630091 | 831411 | 730751 |
| 1980 | 187999 | 31132 | 70739 | 186367 | 228568 | 112192 | 199900 | 550336 | 734489 | 642412 |
| 1981 | 227727 | 31000 | 70441 | 125578 | 151442 | 116284 | 196022 | 527318 | 684352 | 605835 |
| 1982 | 194715 | 23583 | 52338 | 104116 | 125802 | 95513 | 162517 | 439982 | 567499 | 503741 |
| 1983 | 33240 | 17688 | 39712 | 76554 | 94103 | 90339 | 143724 | 236377 | 337388 | 286882 |
| 1984 | 38916 | 13255 | 30019 | 74062 | 88256 | 99798 | 162144 | 245424 | 347471 | 296448 |
| 1985 | 139233 | 18582 | 40002 | 97329 | 118841 | 119727 | 204723 | 399028 | 539102 | 469065 |
| 1986 | 171745 | 23343 | 50988 | 121610 | 150859 | 94602 | 166818 | 435090 | 575673 | 505381 |
| 1987 | 173687 | 29639 | 65127 | 74996 | 92205 | 100297 | 167564 | 398168 | 527764 | 462966 |
| 1988 | 116767 | 20709 | 44860 | 75300 | 92364 | 86819 | 143371 | 317609 | 423746 | 370678 |
| 1989 | 60693 | 18139 | 39691 | 53173 | 65040 | 92671 | 155576 | 241044 | 345930 | 293487 |
| 1990 | 73109 | 11072 | 24518 | 37739 | 45590 | 82687 | 132871 | 218191 | 296332 | 257262 |
| 1991 | 110680 | 9302 | 20175 | 22639 | 29107 | 93880 | 167425 | 249798 | 349917 | 299857 |
| 1992 | 41855 | 2748 | 6790 | 11967 | 15386 | 77128 | 134924 | 143925 | 216262 | 180094 |
| 1993 | 0 | 1878 | 4441 | 10764 | 13839 | 73080 | 143058 | 95352 | 179428 | 137390 |
| 1994 | 0 | 1018 | 2651 | 7823 | 10058 | 90767 | 183998 | 110985 | 219159 | 165072 |
| 1995 | 21341 | 910 | 2267 | 5090 | 6545 | 82996 | 154188 | 120523 | 202958 | 161740 |
| 1996 | 21944 | 858 | 2006 | 4860 | 6249 | 68533 | 118542 | 104675 | 163182 | 133928 |
| 1997 | 16814 | 1045 | 2367 | 1588 | 2269 | 42131 | 71205 | 69083 | 123311 | 96197 |
| 1998 | 3026 | 161 | 367 | 759 | 1084 | 46654 | 85558 | 58751 | 126207 | 92479 |
| 1999 | 5374 | 142 | 306 | 1082 | 1545 | 43536 | 86471 | 57800 | 130436 | 94118 |
| 2000 | 5571 | 300 | 631 | 0 | 0 | 0 | 0 | 5871 | 6202 | 6036 |

Table 4.2.3.4 Run reconstruction data inputs and estimated pre-fishery abundance for maturing (MN1) 1SW salmon (grilse) of North American origin (terms defined in Table 4.2.3.2).

| $\begin{gathered} 1 \mathrm{SW} \\ \text { Year (i) } \\ \hline \end{gathered}$ | MC1 <br> min (i) | max <br> (i) | MR1 <br> min <br> (i) | max <br> (i) | MN1 <br> min <br> (i) | max <br> (i) | midpoint (i) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 213987 | 267720 | 205245 | 441495 | 421294 | 713652 | 567473 |
| 1972 | 237286 | 279064 | 198169 | 415122 | 437446 | 698358 | 567902 |
| 1973 | 346109 | 408260 | 224721 | 435169 | 573089 | 847803 | 710446 |
| 1974 | 322772 | 379370 | 221491 | 449026 | 546489 | 832909 | 689699 |
| 1975 | 351015 | 422105 | 268639 | 578367 | 622355 | 1006285 | 814320 |
| 1976 | 313060 | 375300 | 299958 | 603740 | 616033 | 985108 | 800571 |
| 1977 | 252058 | 318032 | 223971 | 469268 | 478280 | 792016 | 635148 |
| 1978 | 132546 | 172340 | 169121 | 339201 | 303367 | 514951 | 409159 |
| 1979 | 218442 | 252711 | 233025 | 467126 | 453809 | 724532 | 589171 |
| 1980 | 343344 | 412617 | 296958 | 617145 | 643287 | 1035964 | 839625 |
| 1981 | 308670 | 377651 | 362877 | 762381 | 675194 | 1147695 | 911444 |
| 1982 | 265678 | 312538 | 307129 | 634111 | 575893 | 953022 | 764458 |
| 1983 | 197184 | 234389 | 192838 | 398250 | 391960 | 636641 | 514300 |
| 1984 | 158852 | 187900 | 230933 | 447980 | 392105 | 640382 | 516244 |
| 1985 | 227928 | 259284 | 258290 | 519507 | 488814 | 784012 | 636413 |
| 1986 | 278654 | 321357 | 339768 | 677807 | 621836 | 1005976 | 813906 |
| 1987 | 319510 | 375472 | 328770 | 674567 | 651584 | 1056819 | 854201 |
| 1988 | 240291 | 276488 | 374635 | 750007 | 618691 | 1034033 | 826362 |
| 1989 | 205998 | 239495 | 231101 | 454400 | 439422 | 698461 | 568941 |
| 1990 | 134630 | 156382 | 273654 | 530664 | 411034 | 692380 | 551707 |
| 1991 | 117141 | 133509 | 188679 | 362619 | 307716 | 499773 | 403745 |
| 1992 | 21986 | 30556 | 342594 | 618473 | 368023 | 655245 | 511634 |
| 1993 | 15027 | 19983 | 297095 | 566627 | 315107 | 592304 | 453706 |
| 1994 | 8142 | 11928 | 208098 | 418763 | 218331 | 434899 | 326615 |
| 1995 | 7278 | 10200 | 292082 | 662880 | 302296 | 679741 | 491019 |
| 1996 | 6861 | 9028 | 414848 | 891504 | 425878 | 909491 | 667685 |
| 1997 | 8358 | 10652 | 225194 | 400374 | 235816 | 415050 | 325433 |
| 1998 | 3054 | 3302 | 225603 | 376598 | 240161 | 608110 | 424136 |
| 1999 | 2705 | 2758 | 222221 | 331819 | 236245 | 535653 | 385949 |
| 2000 | 5697 | 5675 | 192865 | 370647 | 208520 | 600927 | 404724 |

Table 4.2.4.1 Estimated numbers of 2SW spawners in North America by geographic regions, 1971-2000.

| Year | Labrador |  | Newfoundland |  | Quebec |  | Gulf of St. Lawrence |  | Scotia-Fundy |  | USA | North America |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max |  | Min | Max | Mid-points |
| 1971 | 4,012 | 28,882 | 1,817 | 8,055 | 11,822 | 17,733 | 4,303 | 8,237 | 4,496 | 9,032 | 490 | 26,940 | 72,429 | 49,684 |
| 1972 | 3,435 | 24,812 | 2,008 | 8,240 | 23,160 | 34,741 | 17,803 | 32,996 | 7,459 | 12,699 | 1,038 | 54,903 | 114,525 | 84,714 |
| 1973 | 4,565 | 34,376 | 2,283 | 10,449 | 23,564 | 35,346 | 20,505 | 38,126 | 3,949 | 7,844 | 1,100 | 55,966 | 127,240 | 91,603 |
| 1974 | 4,490 | 33,475 | 1,510 | 5,942 | 28,657 | 42,985 | 31,702 | 57,923 | 9,526 | 15,979 | 1,147 | 77,032 | 157,451 | 117,242 |
| 1975 | 4,564 | 32,119 | 1,888 | 7,086 | 23,818 | 35,726 | 18,477 | 33,210 | 11,861 | 18,830 | 1,942 | 62,549 | 128,913 | 95,731 |
| 1976 | 4,984 | 36,701 | 2,011 | 7,198 | 22,653 | 33,980 | 14,821 | 29,694 | 11,045 | 18,337 | 1,126 | 56,641 | 127,035 | 91,838 |
| 1977 | 4,042 | 31,969 | 1,114 | 5,088 | 32,602 | 48,902 | 32,535 | 60,188 | 13,578 | 23,119 | 643 | 84,512 | 169,909 | 127,211 |
| 1978 | 3,361 | 25,490 | 1,557 | 5,712 | 29,889 | 44,834 | 11,511 | 22,829 | 6,517 | 11,428 | 3,314 | 56,150 | 113,608 | 84,879 |
| 1979 | 1,823 | 14,528 | 980 | 3,463 | 12,807 | 19,210 | 3,575 | 6,823 | 4,683 | 8,234 | 1,509 | 25,376 | 53,767 | 39,572 |
| 1980 | 4,633 | 34,525 | 1,888 | 6,925 | 35,594 | 53,390 | 19,947 | 37,645 | 14,270 | 25,628 | 4,263 | 80,596 | 162,375 | 121,486 |
| 1981 | 4,403 | 31,615 | 3,074 | 11,442 | 26,132 | 39,199 | 4,657 | 10,028 | 5,870 | 13,353 | 4,334 | 48,470 | 109,971 | 79,221 |
| 1982 | 3,081 | 23,127 | 2,579 | 8,481 | 26,492 | 39,738 | 11,036 | 20,330 | 5,656 | 11,335 | 4,643 | 53,486 | 107,655 | 80,571 |
| 1983 | 2,267 | 16,824 | 2,244 | 7,677 | 17,308 | 25,963 | 7,436 | 14,288 | 1,505 | 6,529 | 1,769 | 32,529 | 73,050 | 52,790 |
| 1984 | 1,478 | 11,822 | 2,063 | 6,800 | 22,345 | 32,659 | 15,332 | 27,195 | 14,245 | 23,650 | 2,547 | 58,011 | 104,673 | 81,342 |
| 1985 | 1,258 | 9,530 | 946 | 3,042 | 20,668 | 31,742 | 21,168 | 39,982 | 18,185 | 33,580 | 4,884 | 67,108 | 122,759 | 94,934 |
| 1986 | 2,177 | 16,334 | 1,575 | 5,198 | 24,088 | 35,939 | 32,991 | 64,980 | 15,435 | 30,120 | 5,570 | 81,836 | 158,141 | 119,988 |
| 1987 | 2,895 | 21,821 | 1,320 | 4,409 | 21,723 | 31,727 | 19,877 | 43,120 | 10,235 | 19,233 | 2,781 | 58,831 | 123,091 | 90,961 |
| 1988 | 1,625 | 13,452 | 1,540 | 5,033 | 25,390 | 38,343 | 23,392 | 44,859 | 9,074 | 18,381 | 3,038 | 64,059 | 123,106 | 93,582 |
| 1989 | 1,727 | 13,270 | 690 | 2,289 | 25,016 | 35,905 | 14,758 | 30,866 | 11,689 | 21,539 | 2,800 | 56,680 | 106,668 | 81,674 |
| 1990 | 923 | 7,493 | 1,327 | 4,372 | 24,422 | 36,219 | 22,554 | 49,478 | 9,688 | 18,245 | 4,356 | 63,269 | 120,163 | 91,716 |
| 1991 | 491 | 3,665 | 1,041 | 3,410 | 19,959 | 29,052 | 19,590 | 41,956 | 9,356 | 16,479 | 2,416 | 52,854 | 96,978 | 74,916 |
| 1992 | 2,012 | 14,889 | 3,057 | 10,474 | 19,337 | 28,833 | 27,448 | 54,168 | 8,725 | 15,280 | 2,292 | 62,871 | 125,936 | 94,403 |
| 1993 | 3,624 | 17,922 | 1,449 | 5,017 | 15,774 | 21,428 | 25,218 | 46,308 | 5,710 | 9,921 | 2,065 | 53,839 | 102,661 | 78,250 |
| 1994 | 5,339 | 23,981 | 1,840 | 6,077 | 15,631 | 21,147 | 20,315 | 54,101 | 3,682 | 6,093 | 1,344 | 48,152 | 112,743 | 80,447 |
| 1995 | 12,006 | 43,726 | 3,563 | 12,481 | 22,575 | 28,703 | 22,634 | 60,511 | 4,672 | 7,971 | 1,748 | 67,198 | 155,140 | 111,169 |
| 1996 | 8,838 | 32,395 | 4,372 | 14,028 | 19,010 | 25,421 | 18,416 | 39,757 | 6,507 | 11,242 | 2,407 | 59,550 | 125,250 | 92,400 |
| 1997 | 9,221 | 23,646 | 3,780 | 8,190 | 15,531 | 20,780 | 15,832 | 35,144 | 3,095 | 5,311 | 1,611 | 49,070 | 94,682 | 71,876 |
| 1998 | - | - | 5,222 | 12,295 | 14,176 | 19,333 | 6,568 | 16,742 | 2,424 | 5,663 | 1,526 | - | - | - |
| 1999 | - | - | 4,169 | 14,126 | 17,198 | 23,723 | 11,372 | 27,406 | 3,041 | 6,648 | 1,168 | - | - | - |
| 2000 | - | - | 1,975 | 16,021 | 16,067 | 23,365 | 12,262 | 27,339 | 1,855 | 4,877 | 533 | - | - | - |

Labrador: SFAs 1,2\&14B
Newfoundland: SFAs 3-14A
Gulf of St. Lawrence: SFAs 15-18
Scotia-Fundy: SFAs 19-23 (SFA 22 is not included as it does not produce 2SW salmon)
Quebec: Q1-Q11

Table 4.2.4.2 Estimated numbers of 1SW spawners in North America by geographic regions, 1971-2000.

|  | Labrador |  | Newfoundland |  | Quebec |  | Gulf of St. Lawrence |  | Scotia-Fundy |  | USA | North America |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max |  | Min | Max | Mid-points |
| 1971 | 29,032 | 111,448 | 85,978 | 199,463 | 9,338 | 14,007 | 19,874 | 35,534 | 4,800 | 12,810 | 29 | 149,051 | 373,291 | 261,171 |
| 1972 | 21,728 | 83,415 | 84,880 | 195,010 | 8,213 | 12,320 | 24,319 | 43,318 | 2,992 | 10,385 | 17 | 142,149 | 344,465 | 243,307 |
| 1973 | 0 | 11,405 | 108,785 | 253,965 | 10,987 | 16,480 | 28,105 | 51,257 | 8,658 | 18,715 | 13 | 156,548 | 351,834 | 254,191 |
| 1974 | 24,533 | 92,118 | 58,731 | 144,263 | 10,067 | 15,100 | 48,343 | 84,685 | 16,209 | 33,822 | 40 | 157,922 | 370,028 | 263,975 |
| 1975 | 49,688 | 183,837 | 78,882 | 191,775 | 11,606 | 17,409 | 42,668 | 74,920 | 18,232 | 28,608 | 67 | 201,143 | 496,615 | 348,879 |
| 1976 | 31,814 | 125,665 | 80,571 | 196,132 | 12,979 | 19,469 | 56,021 | 99,810 | 24,589 | 43,595 | 151 | 206,125 | 484,822 | 345,474 |
| 1977 | 28,815 | 112,337 | 75,762 | 186,149 | 12,004 | 18,006 | 14,045 | 27,585 | 16,704 | 34,231 | 54 | 147,385 | 378,364 | 262,874 |
| 1978 | 13,464 | 53,851 | 68,756 | 166,429 | 11,447 | 17,170 | 13,768 | 25,474 | 5,678 | 9,808 | 127 | 113,240 | 272,859 | 193,049 |
| 1979 | 17,825 | 72,682 | 76,233 | 183,991 | 15,863 | 23,795 | 29,764 | 57,382 | 18,577 | 36,754 | 247 | 158,508 | 374,850 | 266,679 |
| 1980 | 45,870 | 170,045 | 85,189 | 206,650 | 20,817 | 31,226 | 26,450 | 50,297 | 28,878 | 52,513 | 722 | 207,926 | 511,453 | 359,690 |
| 1981 | 49,855 | 187,471 | 110,755 | 268,677 | 30,952 | 46,428 | 39,421 | 77,501 | 18,236 | 42,948 | 1,009 | 250,228 | 624,035 | 437,132 |
| 1982 | 34,032 | 129,370 | 99,376 | 241,131 | 16,877 | 25,316 | 52,020 | 97,071 | 12,179 | 26,548 | 290 | 214,774 | 519,727 | 367,250 |
| 1983 | 19,360 | 78,689 | 77,514 | 187,796 | 12,030 | 18,045 | 13,611 | 24,683 | 7,747 | 15,969 | 255 | 130,517 | 325,436 | 227,976 |
| 1984 | 9,348 | 40,056 | 91,505 | 222,730 | 16,316 | 24,957 | 17,990 | 33,657 | 17,964 | 37,503 | 540 | 153,663 | 359,444 | 256,554 |
| 1985 | 19,631 | 76,462 | 85,179 | 207,175 | 15,608 | 25,140 | 39,514 | 73,906 | 18,158 | 40,731 | 363 | 178,454 | 423,778 | 301,116 |
| 1986 | 30,806 | 116,481 | 87,833 | 213,537 | 22,230 | 33,855 | 82,122 | 149,587 | 21,204 | 44,947 | 660 | 244,854 | 559,067 | 401,960 |
| 1987 | 37,572 | 144,917 | 104,096 | 232,991 | 25,789 | 40,481 | 59,330 | 110,335 | 21,589 | 45,407 | 1,087 | 249,463 | 575,217 | 412,340 |
| 1988 | 34,369 | 134,100 | 93,396 | 227,054 | 28,582 | 44,815 | 85,644 | 159,916 | 23,288 | 47,231 | 923 | 266,203 | 614,039 | 440,121 |
| 1989 | 22,429 | 90,212 | 41,798 | 102,199 | 24,710 | 37,319 | 44,715 | 81,719 | 23,873 | 48,578 | 1,080 | 158,605 | 361,108 | 259,857 |
| 1990 | 12,544 | 52,176 | 69,576 | 169,449 | 26,594 | 39,826 | 56,161 | 113,442 | 22,753 | 49,642 | 617 | 188,245 | 425,153 | 306,699 |
| 1991 | 10,526 | 42,647 | 44,023 | 108,779 | 20,582 | 30,433 | 44,350 | 87,876 | 13,814 | 25,610 | 235 | 133,530 | 295,580 | 214,555 |
| 1992 | 15,229 | 59,331 | 95,096 | 214,129 | 21,754 | 33,583 | 118,723 | 189,260 | 15,125 | 29,633 | 1,124 | 267,051 | 527,060 | 397,056 |
| 1993 | 22,499 | 78,251 | 107,816 | 242,217 | 17,493 | 27,444 | 70,969 | 118,119 | 11,539 | 22,252 | 444 | 230,760 | 488,726 | 359,743 |
| 1994 | 15,228 | 53,958 | 66,185 | 162,342 | 16,758 | 25,642 | 32,651 | 90,339 | 6,918 | 10,218 | 427 | 138,167 | 342,925 | 240,546 |
| 1995 | 22,144 | 73,575 | 172,727 | 405,141 | 14,409 | 21,548 | 15,407 | 61,251 | 12,114 | 19,697 | 213 | 237,014 | 581,424 | 409,219 |
| 1996 | 48,362 | 150,048 | 218,639 | 520,504 | 18,923 | 27,805 | 24,411 | 70,260 | 19,253 | 32,472 | 651 | 330,240 | 801,740 | 565,990 |
| 1997 | 64,049 | 153,200 | 80,096 | 127,116 | 14,724 | 22,210 | 12,699 | 36,748 | 6,143 | 9,428 | 365 | 178,076 | 349,068 | 263,572 |
| 1998 | - | - | 124,551 | 225,216 | 16,277 | 24,954 | 23,580 | 46,609 | 16,342 | 26,028 | 403 | - | - |  |
| 1999 | - | - | 135,561 | 203,780 | 18,785 | 28,502 | 18,212 | 36,304 | 10,177 | 16,516 | 419 | - | - |  |
| 2000 | - | - | 99,743 | 234,572 | 17,137 | 27,096 | 25,968 | 43,558 | 10,656 | 17,977 | 270 | - | - |  |

Labrador: SFAs $1,2 \& 14 B$
Newfoundland: SFAs 3-14A
Gulf of St. Lawrence: SFAs 15-18
Scotia-Fundy: SFAs 19-23 (SFA 22 is not included as it does not produce 2SW salmon)
Quebec: Q1-Q11

Table 4.2.4.3. Smolt age distributions in six stock areas of North America used to weight forward the spawning escapement in the current year to the year of the non-maturing 1SW component in the Northwest Atlantic.

|  | Smolt age (years) |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Stock area | 1 | 2 | 3 | 4 | 5 | 6 |
| Labrador | 0.0 | 0.0 | 0.077 | 0.542 | 0.341 | 0.040 |
| Newfoundland | 0.0 | 0.041 | 0.598 | 0.324 | 0.038 | 0.0 |
| Québec | 0.0 | 0.058 | 0.464 | 0.378 | 0.089 | 0.010 |
| Gulf of | St. | 0.0 | 0.398 | 0.573 | 0.029 | 0.0 |
| Lawrence |  |  |  |  |  | 0.0 |
| Scotia-Fundy | 0.0 | 0.600 | 0.394 | 0.006 | 0.0 | 0.0 |
| USA | 0.377 | 0.520 | 0.103 | 0.0 | 0.0 | 0.0 |

Table 4.2.4.4 The mid-point of 2 SW spawners and lagged spawners for North America and to each of the geographic areas. Lagged refers to the allocation of spawners to the year
in which they would have contributed to the year of prefishery abundance.

| Year | North America |  | Prefishery Recruits/ abundance 2SW lagged recruits spawner |  | Labrador (L) |  | Newfoundland ( N ) |  | Quebec (Q) |  | Gulf of St. Lawrence (G) |  | Scotia-Fundy (S) |  | USA (US) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Total 2SW spawners | Lagged 2SW spawners |  |  | Total | Lagged | Total | Lagged | Total | Lagged | Total | Lagged | Total | Lagged | Total | Lagged |
| 1971 | 49684 |  | 652827 |  | 16447 |  | 4936 |  | 14777 |  | 6270 |  | 6764 |  | 490 |  |
| 1972 | 84714 |  | 645486 |  | 14124 |  | 5124 |  | 28951 |  | 25399 |  | 10079 |  | 1038 |  |
| 1973 | 91603 |  | 770200 |  | 19470 |  | 6366 |  | 29455 |  | 29316 |  | 5896 |  | 1100 |  |
| 1974 | 117242 |  | 712403 |  | 18982 |  | 3726 |  | 35821 |  | 44813 |  | 12752 |  | 1147 |  |
| 1975 | 95731 |  | 807391 |  | 18341 |  | 4487 |  | 29772 |  | 25844 |  | 15345 |  | 1942 |  |
| 1976 | 91838 |  | 718805 |  | 20842 |  | 4605 |  | 28316 |  | 22258 |  | 14691 |  | 1126 |  |
| 1977 | 127211 |  | 587326 |  | 18006 |  | 3101 |  | 40752 |  | 46361 |  | 18348 |  | 643 |  |
| 1978 | 84879 | 95423 | 330077 | 3.46 | 14425 | 14759 | 3635 | 5802 | 37362 | 28016 | 17170 | 35371 | 8973 | 10034 | 3314 | 1442 |
| 1979 | 39572 | 107023 | 730725 | 6.83 | 8175 | 17486 | 2221 | 4664 | 16008 | 32232 | 5199 | 36818 | 6459 | 14270 | 1509 | 1553 |
| 1980 | 121486 | 96095 | 639192 | 6.65 | 19579 | 18903 | 4406 | 4316 | 44492 | 31940 | 28796 | 24971 | 19949 | 14937 | 4263 | 1029 |
| 1981 | 79221 | 104076 | 605935 | 5.82 | 18009 | 18795 | 7258 | 4472 | 32666 | 30266 | 7342 | 31955 | 9612 | 16888 | 4334 | 1699 |
| 1982 | 80571 | 107284 | 503481 | 4.69 | 13104 | 19695 | 5530 | 3661 | 33115 | 34821 | 15683 | 34049 | 8496 | 12699 | 4643 | 2358 |
| 1983 | 52790 | 82182 | 286891 | 3.49 | 9546 | 18710 | 4961 | 3440 | 21636 | 36526 | 10862 | 13258 | 4017 | 7514 | 1769 | 2733 |
| 1984 | 81342 | 79799 | 296450 | 3.71 | 6650 | 15422 | 4432 | 2801 | 27502 | 28065 | 21264 | 14937 | 18947 | 14569 | 2547 | 4006 |
| 1985 | 94934 | 85408 | 468776 | 5.49 | 5394 | 11576 | 1994 | 3786 | 26205 | 32359 | 30575 | 19576 | 25882 | 13668 | 4884 | 4443 |
| 1986 | 119988 | 80977 | 505066 | 6.24 | 9255 | 15361 | 3386 | 6075 | 30013 | 35728 | 48985 | 11286 | 22777 | 8998 | 5570 | 3528 |
| 1987 | 90961 | 78610 | 462953 | 5.89 | 12358 | 17772 | 2865 | 6023 | 26725 | 33119 | 31498 | 13524 | 14734 | 5813 | 2781 | 2359 |
| 1988 | 93582 | 79001 | 370526 | 4.69 | 7538 | 14762 | 3287 | 5209 | 31866 | 27538 | 34125 | 15142 | 13728 | 13002 | 3038 | 3347 |
| 1989 | 81674 | 93776 | 293057 | 3.13 | 7498 | 10875 | 1490 | 4544 | 30461 | 25762 | 22812 | 24668 | 16614 | 23026 | 2800 | 4901 |
| 1990 | 91716 | 103388 | 256969 | 2.49 | 4208 | 7799 | 2850 | 2951 | 30320 | 26580 | 36016 | 37632 | 13966 | 23978 | 4356 | 4449 |
| 1991 | 74916 | 99937 | 299086 | 2.99 | 2078 | 6285 | 2225 | 2953 | 24506 | 28072 | 30773 | 41497 | 12917 | 17965 | 2416 | 3166 |
| 1992 | 94403 | 89467 | 179755 | 2.01 | 8451 | 8072 | 6765 | 3018 | 24085 | 28227 | 40808 | 33056 | 12002 | 14173 | 2292 | 2922 |
| 1993 | 78250 | 91771 | 137134 | 1.49 | 10773 | 10649 | 3233 | 3080 | 18601 | 29616 | 35763 | 29551 | 7816 | 15464 | 2065 | 3410 |
| 1994 | 80447 | 88940 | 161214 | 1.81 | 14660 | 9247 | 3958 | 2178 | 18389 | 30646 | 37208 | 28397 | 4888 | 15007 | 1344 | 3464 |
| 1995 | 11169 | 89461 | 156490 | 1.75 | 27866 | 7453 | 8022 | 2400 | 25639 | 30138 | 41572 | 33549 | 6322 | 13350 | 1748 | 2570 |
| 1996 | 92400 | 84687 | 126588 | 1.49 | 20617 | 5299 | 9200 | 2585 | 22216 | 27289 | 29086 | 34922 | 8875 | 12373 | 2407 | 2219 |
| 1997 | 71876 | 82888 | 97899 | 1.18 | 16434 | 3511 | 5985 | 5004 | 18155 | 24550 | 25488 | 38513 | 4203 | 9493 | 1611 | 1817 |
| 1998 |  | 76104 |  |  |  | 6285 | 8758 | 4368 | 16754 | 21312 | 11655 | 36488 | 4044 | 6080 | 1526 | 1571 |
| 1999 |  | 80008 |  |  |  | 9930 | 9148 | 3994 | 20460 | 19459 | 19389 | 38906 | 4845 | 5764 | 1168 | 1954 |
| 2000 |  | 89093 |  |  |  | 14098 |  | 6574 |  | 22055 |  | 36481 |  | 7845 |  | 2039 |
| 2001 |  | 89243 |  |  |  | 22118 |  | 8490 |  | 22898 |  | 28021 |  | 6056 |  | 1661 |
| 2002 |  | 75646 |  |  |  | 22527 |  | 7215 |  | 20281 |  | 20091 |  | 4133 |  | 1400 |
| 2003 |  |  |  |  |  |  |  | 7892 |  | 18078 |  | 15136 |  | 4525 |  | 2908 |

Spawners lagged by:
Labrador $=0.0768 \times \mathrm{i}-5$ spawners $+0.542 \times \mathrm{i}-6+0.341 \times \mathrm{i}-7+0.0401 \times \mathrm{i}-8$
Newfoundland $=0.0408 \times i-4$ spawners $+0.5979 \times i-5+0.3237 \times \mathrm{i}-6+0.0375 \times \mathrm{i}-7$
Newfoundland $=0.0408 \times i-4$ spawners $+0.5979 \times i-5+0.3237 \times i-6+0.0375 \times i-7$
Quebec $=0.0577 \times i-4$ spawners $+0.4644 \times \mathrm{i}-5+0.3783 \times i-6+0.0892 \times i-7+0.0104 \times i-8$
Gulf $=0.3979 \times i-4$ spawners $+0.5731 \times i-5+0.0291 \times i-6$
Scotia-Fundy $=0.6002 \times \mathrm{i}-4$ spawners $+0.3942 \times \mathrm{i}-5+0.0055 \times \mathrm{i}-6$
USA $=0.3767 \times i-3$ spawners $+0.520 \times i-4+0.1033 \times i-5$.

Table 4.4.1. 2SW spawning requirements for North America by country, management zone and overall. Management zones are shown in Figure 4.1.1.1.

| Country | Stock Area | Management zone | 2SW spawner requirement |  |
| :---: | :---: | :---: | :---: | :---: |
| Canada | Labrador | SFA 1 | 7,992 |  |
|  |  | SFA 2 | 25,369 |  |
|  |  | SFA 14B | 1,390 |  |
|  | Subtotal |  |  | 34,746 |
|  | Newfoundland | SFA 3 | 240 |  |
|  |  | SFA 4 | 488 |  |
|  |  | SFA 5 | 233 |  |
|  |  | SFA 6 to 8 | 13 |  |
|  |  | SFA 9 to 12 | 212 |  |
|  |  | SFA 13 | 2,544 |  |
|  |  | SFA 14A | 292 |  |
|  | Subtotal |  |  | 4,022 |
|  | Gulf of St. Lawrence | SFA 15 | 5,656 |  |
|  |  | SFA 16 | 21,050 |  |
|  |  | SFA 17 | 537 |  |
|  |  | SFA 18 | 3,187 |  |
|  | Subtotal |  |  | 30,430 |
|  | Québec | Q1 | 2,532 |  |
|  |  | Q2 | 1,797 |  |
|  |  | Q3 | 1,788 |  |
|  |  | Q5 | 948 |  |
|  |  | Q6 | 818 |  |
|  |  | Q7 | 2,021 |  |
|  |  | Q8 | 11,195 |  |
|  |  | Q9 | 3,378 |  |
|  |  | Q10 | 1,582 |  |
|  |  | Q11 | 3,387 |  |
|  | Subtotal |  | 29,446 |  |
|  | Scotia-Fundy | SFA 19 | 3,138 |  |
|  |  | SFA 20 | 2,691 |  |
|  |  | SFA 21 | 5,817 |  |
|  |  | SFA 22 | 0 |  |
|  |  | SFA 23 | 13,059 |  |
|  | Subtotal |  | 24,705 |  |
| Total |  |  | 123,349 |  |
| USA | Connecticut <br> Merrimack |  | 9,727 |  |
|  |  |  | 2,599 |  |
|  | Penobscot |  | 6,838 |  |
|  | Other Maine rivers |  | 9,668 |  |
|  | Paucatuck |  | 367 |  |
| Total |  |  | 29,199 |  |
| North American Total |  |  | 152,548 |  |

Table 4.5.1 Fishing mortalities of 2SW salmon equivalents by North American fisheries, 1972-2000.
Only mid-points of the estimated values have been used.

| Year | CANADA |  |  |  |  |  |  |  |  |  | USA | Total | Terminal Fisheries as a $\%$ of Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | MIXED STOCK |  |  |  | TERMINAL FISHERIES IN YEAR i |  |  |  |  |  |  |  |  |
|  | NF-LAB <br> Comm 1SW <br> (Yr i-1) <br> (b) | \% 1SW of total 2SW equivalents | $\begin{gathered} \text { NF-LAB } \\ \text { Comm 2SW } \\ \text { (Yri) (b) } \\ \hline \end{gathered}$ | $\begin{gathered} \text { NF-Lab } \\ \text { comm total } \end{gathered}$ | Labrador rivers (a) | Nfld rivers <br> (a) | Quebec <br> Region | $\begin{array}{r} \text { Gulf } \\ \text { Region } \\ \hline \end{array}$ | Scotia - <br> Fundy <br> Region | Canadian total | Year i |  |  |
| 1972 | 27,874 | 11 | 156,881 | 184,755 | 314 | 633 | 27,417 | 22,389 | 6,801 | 242,310 | 346 | 242,656 | 24 |
| 1973 | 24,016 | 8 | 223,603 | 247,619 | 719 | 895 | 32,751 | 17,915 | 6,680 | 306,580 | 327 | 306,907 | 19 |
| 1974 | 32,828 | 9 | 240,676 | 273,504 | 593 | 542 | 47,631 | 21,429 | 12,734 | 356,434 | 247 | 356,681 | 23 |
| 1975 | 32,316 | 9 | 242,398 | 274,714 | 241 | 528 | 41,097 | 15,675 | 12,375 | 344,629 | 389 | 345,018 | 20 |
| 1976 | 47,846 | 13 | 261,770 | 309,616 | 618 | 412 | 42,139 | 18,088 | 11,111 | 381,985 | 191 | 382,176 | 19 |
| 1977 | 36,777 | 10 | 246,090 | 282,867 | 954 | 946 | 42,301 | 33,433 | 15,562 | 376,062 | 1,355 | 377,418 | 25 |
| 1978 | 37,200 | 14 | 160,477 | 197,677 | 580 | 559 | 37,421 | 23,803 | 10,781 | 270,821 | 894 | 271,714 | 27 |
| 1979 | 18,825 | 13 | 93,917 | 112,742 | 469 | 144 | 25,234 | 6,299 | 4,506 | 149,395 | 433 | 149,828 | 25 |
| 1980 | 27,923 | 8 | 221,597 | 249,520 | 646 | 699 | 53,567 | 29,828 | 18,411 | 352,670 | 1,533 | 354,202 | 30 |
| 1981 | 46,088 | 14 | 205,403 | 251,492 | 384 | 485 | 44,375 | 16,326 | 13,988 | 327,050 | 1,267 | 328,317 | 23 |
| 1982 | 45,894 | 18 | 137,132 | 183,026 | 473 | 433 | 35,204 | 25,707 | 12,353 | 257,195 | 1,413 | 258,608 | 29 |
| 1983 | 34,348 | 15 | 113,815 | 148,163 | 313 | 445 | 34,472 | 27,094 | 13,515 | 224,002 | 386 | 224,388 | 34 |
| 1984 | 25,969 | 18 | 84,480 | 110,448 | 379 | 215 | 24,408 | 6,041 | 3,971 | 145,464 | 675 | 146,138 | 24 |
| 1985 | 19,578 | 14 | 80,351 | 99,929 | 219 | 15 | 27,483 | 2,745 | 4,930 | 135,322 | 645 | 135,967 | 27 |
| 1986 | 26,504 | 15 | 107,009 | 133,514 | 340 | 39 | 33,846 | 4,582 | 2,824 | 175,145 | 606 | 175,750 | 24 |
| 1987 | 33,629 | 16 | 134,879 | 168,508 | 457 | 20 | 33,807 | 3,795 | 1,370 | 207,956 | 300 | 208,256 | 19 |
| 1988 | 42,874 | 26 | 82,769 | 125,642 | 514 | 29 | 34,262 | 3,922 | 1,373 | 165,743 | 248 | 165,990 | 24 |
| 1989 | 29,664 | 20 | 82,998 | 112,662 | 337 | 9 | 28,901 | 3,513 | 265 | 145,686 | 397 | 146,083 | 23 |
| 1990 | 26,164 | 22 | 58,518 | 84,682 | 261 | 24 | 27,986 | 2,847 | 593 | 116,394 | 696 | 117,089 | 28 |
| 1991 | 16,101 | 18 | 41,250 | 57,352 | 66 | 16 | 29,277 | 1,942 | 1,331 | 89,984 | 231 | 90,215 | 36 |
| 1992 | 13,336 | 18 | 25,615 | 38,952 | 581 | 67 | 30,016 | 4,303 | 1,114 | 75,033 | 167 | 75,201 | 48 |
| 1993 | 4,315 | 9 | 13,541 | 17,856 | 273 | 63 | 23,153 | 3,010 | 1,110 | 45,466 | 166 | 45,632 | 61 |
| 1994 | 2,859 | 7 | 12,179 | 15,038 | 365 | 80 | 24,052 | 2,368 | 756 | 42,659 | 1 | 42,660 | 65 |
| $1995$ | 1,660 | 5 | 8,852 | 10,511 | 420 | 92 | 23,331 | 2,041 | 330 | 36,725 | 0 | 36,725 | 71 |
| 1996 | 1,437 | 4 | 5,760 | 7,197 | 320 | 108 | 22,413 | 2,586 | 766 | 33,389 | 0 | 33,389 | 78 |
| 1997 | 1,296 | 5 | 5,499 | 6,795 | 175 | 136 | 18,574 | 2,196 | 581 | 28,456 | 0 | 28,456 | 76 |
| 1998 | 1,544 | 9 | 1,909 | 3,453 | 268 | 129 | 11,256 | 2,224 | 322 | 17,651 | 0 | 17,651 | 80 |
| $1999$ | $239$ | 2 | 912 | 1,151 | 268 | 111 | 9,032 | 1,504 | 450 | 12,515 | 0 | 12,515 | 91 |
| 2000 | 203 | 1 | 1,300 | 1,503 | 268 | 291 | 9,903 | 2,203 | 193 | 14,361 | 0 | 14,361 | 90 |
| 2001 | 421 | - | - | , | - | - | 9, | , |  | - | - | - | - |

NF-Lab comm as $1 \mathrm{SW}=\mathrm{NCl}(\mathrm{mid}-\mathrm{pt}) * 0.904837$
NF-Lab comm as 2SW $=$ NC2 $($ mid-pt $) * 0.99005$
NF-Lab comm as $2 \mathrm{SW}=\mathrm{NC2}(\mathrm{mid}-\mathrm{pt}) * * 0.99005$
Terminal fisheries $=2 \mathrm{SW}$ returns $($ mid-pt $)-2$ SW spawners (mid-pt)
a starting in 1993 , includes estimated mortality of $10 \%$ on hook and released fish
b - starting in 1998, there was no commercial fishery in Labrador; numbers reflect size of aboriginal fish harvest in 1998-2000 and resident food fishery harvest in 2000

Table 4.5.1.1. Catch options for 2001 North American fisheries

| Catch Options for 2001 North American Fisheries (Probability levels refer to <br> probability <br> density function estimates of pre-fishery abundance) |  |  |
| :---: | :---: | :---: |
|  | Pre-fishery Abundance <br> Forecast | Catch Options in 2SW <br> Salmon Equivalents (no.) |
| Probability Level | 145,125 | 0 |
| $\mathbf{2 5}$ | 160,214 | 0 |
| $\mathbf{3 0}$ | 175,591 | 0 |
| $\mathbf{3 5}$ | 191,502 | 14,255 |
| $\mathbf{4 0}$ | 208,016 | 29,127 |
| $\mathbf{4 5}$ | 225,708 | 45,206 |
| $\mathbf{5 0}$ | 244,830 | 62,508 |
| $\mathbf{5 5}$ | 265,996 | 81,660 |
| $\mathbf{6 0}$ | 289,541 | 102,964 |
| $\mathbf{6 5}$ | 316,274 | 127,153 |
| $\mathbf{7 0}$ | 347,994 | 155,855 |
| $\mathbf{7 5}$ |  |  |

Table 4.5.2 History of fishing-related mortalities of North American salmon as 2SW equivalents, 1972-2000.

| Year | Canadian <br> total | USA <br> total | North <br> America <br> Grand <br> Total | \% USA <br> of Total <br> North <br> American | Greenland total | NW <br> Atlantic <br> Total | Harvest in homewaters as \% of total NW Atlantic |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 242,310 | 346 | 242,656 | 0.14 | 260,296 | 502,952 | 48 |
| 1973 | 306,580 | 327 | 306,907 | 0.11 | 181,677 | 488,584 | 63 |
| 1974 | 356,434 | 247 | 356,681 | 0.07 | 218,512 | 575,193 | 62 |
| 1975 | 344,629 | 389 | 345,018 | 0.11 | 199,593 | 544,611 | 63 |
| 1976 | 381,985 | 191 | 382,176 | 0.05 | 252,304 | 634,479 | 60 |
| 1977 | 376,062 | 1,355 | 377,418 | 0.36 | 141,060 | 518,478 | 73 |
| 1978 | 270,821 | 894 | 271,714 | 0.33 | 171,656 | 443,370 | 61 |
| 1979 | 149,395 | 433 | 149,828 | 0.29 | 107,543 | 257,370 | 58 |
| 1980 | 352,670 | 1,533 | 354,202 | 0.43 | 181,023 | 535,225 | 66 |
| 1981 | 327,050 | 1,267 | 328,317 | 0.39 | 170,108 | 498,425 | 66 |
| 1982 | 257,195 | 1,413 | 258,608 | 0.55 | 206,056 | 464,664 | 56 |
| 1983 | 224,002 | 386 | 224,388 | 0.17 | 176,185 | 400,574 | 56 |
| 1984 | 145,464 | 675 | 146,138 | 0.46 | 30,077 | 176,215 | 83 |
| 1985 | 135,322 | 645 | 135,967 | 0.47 | 35,213 | 171,179 | 79 |
| 1986 | 175,145 | 606 | 175,750 | 0.34 | 125,983 | 301,734 | 58 |
| 1987 | 207,956 | 300 | 208,256 | 0.14 | 155,401 | 363,658 | 57 |
| 1988 | 165,743 | 248 | 165,990 | 0.15 | 157,158 | 323,149 | 51 |
| 1989 | 145,686 | 397 | 146,083 | 0.27 | 105,655 | 251,738 | 58 |
| 1990 | 116,394 | 696 | 117,089 | 0.59 | 54,917 | 172,007 | 68 |
| 1991 | 89,984 | 231 | 90,215 | 0.26 | 66,152 | 156,366 | 58 |
| 1992 | 75,033 | 167 | 75,201 | 0.22 | 100,147 | 175,348 | 43 |
| 1993 | 45,466 | 166 | 45,632 | 0.36 | 37,872 | 83,504 | 55 |
| 1994 | 42,659 | 1 | 42,660 | 0.00 | 0 | 42,660 | 100 |
| 1995 | 36,725 | 0 | 36,725 | 0.00 | 0 | 36,725 | 100 |
| 1996 | 33,389 | 0 | 33,389 | 0.00 | 19,310 | 52,699 | 63 |
| 1997 | 28,456 | 0 | 28,456 | 0.00 | 19,856 | 48,312 | 59 |
| 1998 | 17,651 | 0 | 17,651 | 0.00 | 15,214 | 32,865 | 54 |
| 1999 | 12,515 | 0 | 12,515 | 0.00 | 2,738 | 15,253 | 82 |
| 2000 | 14,361 | 0 | 14,361 | 0.00 | 4,863 | 19,223 | 75 |
| 2001 | - | - | - | - | 5,041 | - | - |

Greenland harvest of 2SW equivalents $=$ NG1 $* 0.904837$

Figure 4.1.1.1. Map of Salmon Fishing Areas (SFAs) and Quebec Management Zones (Qs) in Canada.


Figure 4.1.1.2. Summary of recreational fisheries management in eastern Canada and Maine (U.S.A.) during 2000.


Figure 4.1.2.1. Harvest ( t ) of small salmon, large salmon, and combined in Canada, 1960-2000 by all users.


Figure 4.1.2.2. Harvest (number) of small and large salmon and both sizes combined in the recreational fisheries of Canada, 1974 to 2000.


Figure 4.1.2.3. Angling catches (including kept and released fish) of small and large salmon by management area in 2000 (black square) expressed as a proportion of the average catches for the period 1984 to 1991 . The vertical lines represent the minimum to maximum range. The 1984 to 1991 standard period was selected to represent the period of no commercial fisheries in SFAs 15 to 23 and Zones Q1 to Q6 and before the commercial salmon moratorium in Newfoundland SFAs 3 to 14A introduced in 1992. There were no estimates available for releaseed salmon in Newfoundland SFAs 3 to 11 for the years 1984 to 1991. The angling data for SFA 16 in 2000 are not available.



Figure 4.1.2.4. Harvest ( t ) of small salmon and large salmon and both size groups combined in the commercial fisheries of Canada, 1974 to 2000. All commercial fisheries were closed in 2000.


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## REPORT OF THE

# WORKING GROUP ON NORTH ATLANTIC SALMON <br> Part two 

Aberdeen, Scotland

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International Council for the Exploration of the Sea
Conseil International pour l'Exploration de la Mer

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### 5.1 Description of fishery at West Greenland

### 5.1.1 Catch and effort in 2000

At its annual meeting in 1999 the West Greenland Commission of NASCO agreed on a multi-year approach for conservation of the salmon stocks occurring in Greenland. This agreement specified that the catch at West Greenland in 1999 and 2000 should be restricted to that amount used for internal consumption in Greenland, which in the past has been estimated at 20 t . The Greenland authorities subsequently set this amount as total allowable catch.

The fishery was opened on August 14 and was quickly closed by the authorities four days later (August 18) as the reported catch rapidly approached the total allowable catch. The total reported catches in round fresh weight amounted to 20.5 t (Table 5.1.1.1). The geographical distribution of catches by Greenland vessels is given in Table 5.1.1.2 for the years 1977-2000. Compared to previous years, a higher proportion of catch occurred in southern Greenland with $65 \%$ taken in NAFO Division 1F.

By regulation, all catches including landings to local markets, privately purchased salmon, and salmon caught by food fishermen, were reported on a daily basis to the Fishery Licence Office. A private company was given permission to buy salmon from fishermen, and biological samples were purchased from this company.

Licences for the salmon fishery have been issued to fishermen fishing for the local markets, hotels, hospitals etc., while fishing for personal use was permitted without licence for residents of Greenland. In total, 179 licences were issued, however only 45 licensed and 1 unlicensed fishermen reported landings. Due to the arrangement with the purchasing company and the extremely short season only few landings to local markets were observed. Despite the early closure of the fishing season and the considerably increased efficiency of reporting and enforcement system, a relatively large part (one third) of the total fishery is still considered to remain unreported. The unreported catches are estimated to be approximately 10 t in 2000 .

### 5.1.2 Origin of catches at West Greenland

Tissue and biological samples were collected from the mixed population at West Greenland caught for local consumption in 2000. A sample of 491 salmon was purchased from the local purchasing company. In 2000 all samples were obtained through the purchasing company that received most of the landed salmon. The sampled salmon were measured, scales were removed for aging, and gutted weight recorded.

A tissue sample was removed and preserved for DNA analysis. Twelve microsatellite loci were screened in all fish. For Atlantic salmon, these loci have been shown to provide $100 \%$ correct assignment to their continent of origin, and $83 \%$ correct classification to country or province of origin (King et al. 1999).

A total 490 tissue samples were collected for DNA analysis from the following areas, 250 from NAFO Div. 1D, and 241 from NAFO Div. 1F. No precise information on the date of landing could be obtained because samples were taken after landing and transporting of the fish. By assuming August 16 (mid-point of the fishing season) as the landed date, there is a maximum error of $+/-2$ days.

Based on DNA analysis in NAFO Div. 1D $89.2 \%$ of the 250 salmon were of North American origin and $10.8 \%$ of European origin. Of 240 salmon taken in NAFO Div. 1F 50.4 \% were classified as North American while $49.6 \%$ were European. The combined total for the two NAFO divisions sampled is 344 salmon ( $70.2 \%$ ) of North American origin and 146 salmon ( $29.8 \%$ ) of European origin (Table 5.1.2.1, Figure 5.1.2.1). Preliminary attempts to achieve a finer level of resolution of origin indicates that Canadian origin fish dominated the North American component of the catch and Southern European stocks dominated the European component of the catch.

The Working Group noted that the significant increase in proportion of North American origin salmon at West Greenland in 1995-1999 was concordant with the reduction in the number caught (Table 5.1.2.2). This increase is possibly related to the declining number of non-maturing salmon especially in the Southern European countries. The proportion of North American origin salmon estimated for 2000 is similar to that observed during the mid-1980s.

Applying the results of the above analysis to the reported catch indicated that 12.6 t ( 5,100 salmon) of North American origin and 7.6 t (2,700 salmon) of European origin were landed in West Greenland in 2000. Quota reductions have resulted in a reduction in the numbers of North American salmon landed at West Greenland from 1996-1999. The
number of North American salmon remained about the same as that from 1999 to 2000. The number of landed salmon of European origin was similarly reduced from 1995 to 1999, but increased in 2000 due to a higher proportion of European salmon in the Division 1F. The data for 1982 to 2000 (no data for 1993-94) are summarised in Table 5.1.2.2.

### 5.1.3 Biological characteristics of the catches

Biological characteristics (length, weight, and age) were recorded from 491 fish in catches from NAFO Div. 1D and 1F in 2000 and presented in Tables 5.1.3.1 to 5.1.3.3 together with corresponding data from sampling in Greenland since 1968.

The general downward trend in mean length and weight of both European and North American 1SW salmon observed from 1969 - 1995 reversed in 1996 when mean lengths and weights increased (Table 5.1.3.1, Figure 5.1.3.1). From 1996 to 1998 the mean lengths and weights were relatively stable but increased significantly in 1999. In 2000, a decrease was observed, and the mean lengths and weights were among the lowest observed in the time series (Table 5.1.3.1, Figure 5.1.3.1). The Working Group noted that all catches in 2000 were taken relatively early in the traditional fishing season (mid August), i.e. also early in the feeding season in which period the weight increase is known to be $2-3 \%$ per week. The working group noted that the samples in 2000 could be biased towards smaller (younger) fish, as some sorting might have taken place before landing to the purchasing company.

Distribution of the catch by river age in 1968-2000 as determined from scale samples is shown in Table 5.1.3.2. The proportion of the European origin salmon that were river age 1 fish has been quite variable through the later years with relatively high values in 1998-2000 ( $28.6,27.7$ and $36.5 \%$, respectively), the 2000 value being the highest on record. A high proportion of this group suggests a high contribution from Southern European stocks. In 1998 and 1999 low proportions of 7.6 and $7.2 \%$, respectively, of river age 3 were observed, the lowest on record. Some increase from 1999 to 2000 (to $13.1 \%$ ) was observed, close to the overall mean of $16.2 \%$. The proportion of river age 2 salmon of North American origin declined somewhat from 1998, which was close to the overall mean value of $34.8 \%$, to 23.5 and 26.6 \% in 1999 and 2000, respectively.

The sea-age composition of the samples collected from the West Greenland fishery showed no significant changes in the proportions in the North American component of fish from 1998 to 2000 (Table 5.1.3.3), the 1SW proportion being among the highest in the time series. The proportion of 1SW salmon in the European component has been very high since 1997 ( $99.3 \%$ ), and was in 1999 and 2000 estimated at $100 \%$. Samples ( $6 \%$ in weight of the total landings) were representative in time and space of catches in the fishery, but may have been biased if culling of fish occurred prior to purchase for sampling purposes. In addition, the short time period of the fishery may have produced landings that were not representative of the population in the Greenland area in 2000.

### 5.2 Status of the stocks in the West Greenland area

The salmon caught in the West Greenland fishery are non-maturing 1SW or older salmon, nearly all of which would return to home waters in Europe or North America as MSW fish if they survived. While non-maturing 1SW salmon make up more than $90 \%$ of the catch there are also 2 SW salmon and repeat spawners including salmon that had originally spawned for the first time after 1-sea-winter. The most abundant European stocks in West Greenland are thought to originate from the UK and Ireland although low numbers may originate from northern European rivers. For North American MSW salmon, the most abundant stocks in West Greenland are thought to originate in the southern area of the range.

For the Northeast Commission Area, a Run-reconstruction Model was used to update the estimates of pre-fishery abundance of non-maturing 1 SW salmon (Table 3.6.2.4). The main contributor to the abundance of the European component of the West Greenland stock complex is non-maturing 1SW salmon from the southern areas of Europe. These stocks appear to have declined for most of the past 30 years.

Conservation limits and the time series of spawners have been provided for 16 rivers in the NEAC area. Only 6 of the 16 rivers had egg depositions above their conservation limits in the later years. Improvements in egg deposition are indicated for the rivers that were above their conservation limits, while other rivers, which remain near or under their conservation limits on average show a slight decrease (Section 3.4). In general, there seemed to be no indication of recovery from low escapement levels.

In most areas marine survival was lower than the previous 5 -year and 10 -year mean for 1 SW and 2 SW fish. Marine survival rates for 6 hatchery stocks showed a downward trend in survival to home waters for 1 SW and 2 SW salmon for the past 10 -year period.

In general, there has been no significant change in smolt production in the Northeast Atlantic. Returns of salmon to most European rivers showed a significant downward trend for the last ten years period both for southern and northern rivers, but no trend was detected for the last five years.

The Working Group has used the North American run-reconstruction model to estimate pre-fishery abundance, which serves as the basis of abundance forecasts used in the provision of catch advice. The 1999 abundance estimates ranged between 57,800 and 130,436 salmon. The mid-point of this range $(94,118)$ is $2 \%$ higher than the 1998 value $(92,479)$ and is the $2^{\text {nd }}$ lowest in the 30 -year time series (Figure 4.2.3.1). The Working Group expressed concern about the continued decline in the pre-fishery abundance and its impact on spawner levels. While maturing 1SW salmon in 19982000 have increased over the lowest value achieved in 1997, the non-maturing portion of these cohorts has increased only slightly from the lowest value recorded in 1998. Because the pre-fishery abundance has been consistently well below its conservation requirements, this situation is considered to be very serious.

The estimate of the total number of 2SW salmon returning to Newfoundland rivers and coastal waters of other areas of North America in 2000 was close to the number in 1999, but was about $26 \%$ lower than the estimate for 1997 and lower than the average of the previous years (1971-96). The estimates for 1998-2000 are the lowest observed in the past 10 years and among the lowest in the 30 year time series, 1971-2000 (Table 4.2.2.2). The estimates of returns are quite variable over the time series with no trends indicated. Returns have declined from a peak of 226,000 in 1980 to 94,000 in 1999.

In most regions apart from Newfoundland, the returns of 2SW fish in 2000 are near the lower end of the thirty-year time series. However, returns of 2SW salmon to Labrador in 1995 and 1996 were the best in the time series. The estimated returns decreased again in 1997. Closure of the commercial fishery in Labrador after 1997 eliminated the basis for the return and spawner model, and returns could not be estimated for 1998-2000.

The majority of the USA returns were recorded in the rivers of Maine. The estimated 2SW returns and spawners to USA rivers have declined since 1996, and were in $200066 \%$ and $73 \%$ below the previous 5 -year and 10-year averages, respectively. Returns to most USA rivers are hatchery-dependent. Spawning escapements decreased further from $5 \%$ in 1999 to $1.8 \%$ in 2000 compared to conservation requirements.

Egg depositions exceeded or equalled the specific conservation requirements in 23 of the 54 rivers ( $43 \%$ ) assessed in Canada and were less than $50 \%$ of requirements in 18 other rivers ( $33 \%$ ). Large deficiencies in egg depositions were noted in the Bay of Fundy and Atlantic coast of Nova Scotia where 10 of the 11 rivers assessed ( $91 \%$ ) had egg depositions that were less than $50 \%$ of conservation requirements (Figure 4.2.4.1).

North American salmon stocks remain at low levels relative to the 1970s. The 1SW non-maturing component continues to be low with river returns and total production amongst the lowest recorded. The 1999 pre-fishery abundance estimates ranged between 57,800 and 130,436 salmon. The mid-point of this range $(94,118)$ is $2 \%$ higher than the 1998 value $(92,479)$ and is the $2^{\text {nd }}$ lowest in the 28 -year time series. In addition, small salmon returns in 2000 relative to 1999 were generally reduced throughout eastern Canada in the majority of their monitored rivers except in Québec. Returns were similar to or improved ( $>90 \%$ in 2000 relative to 1999) in about half ( $53 \%$ ) of the assessed rivers but this percentage decreased to $45 \%$ without Québec's rivers. Given recent downward trends in returns, improvement in 2SW salmon returns and spawners in 2001 would represent a significant reversal from recently observed trends. Only the Newfoundland stock complex has observed spawning escapements, which have exceeded the area requirements, all other region complexes were below requirement and some declined further in 2000 (Section 4.2.4).

Despite some improvements in the annual returns to some rivers, both in European and North American areas, the overall status of stocks contributing to the West Greenland fishery remains poor, and as a result, the status of stocks within the West Greenland area is thought to be low compared to earlier (historical) levels.

### 5.3 Changes in the continent of origin of salmon captured at West Greenland including changes in migration patterns

The Working Group noted the considerable increase in proportion of North American origin salmon in the fishery at West Greenland in recent years. The proportion of North American origin salmon has changed dramatically over the period of observation, 1969-1999, from below $40 \%$ to a record high level of $90 \%$ in 1999. The proportion of North American origin salmon declined in 2000 fishery samples; however, this may have been due to the early opening and short duration of the fishery. Thus, the catch samples while being descriptive of the fishery may not be a good representation of the salmon population at Greenland. The biological explanation(s) for these changes in North American and European salmon will continue to allude us due to incomplete knowledge of migration of the various components contributing to the West Greenland fishery and more importantly the relative contributions of various stock
groupings. Previous tagging studies including tagging at West Greenland had shown that the southern European stock group contributed more heavily to Greenland than did the northern group. Within North America, it has been shown that stocks in the Gulf of St. Lawrence contributed more heavily than others to Greenland. The DNA analysis in 2000 showed that that annual variations in proportional contributions do occur. Exploratory work into more detailed discrimination of origin of salmon captured at West Greenland will lead to a greater understanding of the mixed stock fishery.

The Working Group examined an analysis of the North American proportions from 1987 to 1999. The year 2000 samples were not included because of the short time scale and geographical distribution of the catch and samples.

## Analysis of Variance for North American proportion at West Greenland



The results of the analysis of variance indicate that the North American proportion varies over year, between NAFO Divisions and that there is a significant interaction effect between year and the various NAFO Divisions. In terms of NAFO Divisions, the North American proportion increased from NAFO Div. 1B to 1C then declined from 1D to 1E and 1F (Figure 5.3.1). The North American proportion has increased significantly from 1987 to 1999. The reasons for the varying North American proportions between NAFO Divisions and years is not completely known. However, it possibly results from different migration patterns and arrival at Greenland of the various stock components.

To learn more about the reasons behind the increasing North American proportion in Greenland a new variable was created by summing the pre-fishery abundances of North American and European non-maturing Atlantic salmon. Examination of the trends in North American proportion at Greenland and in the total pre-fishery abundance of North American plus southern European salmon indicates that the latter is actually declining ( $\mathrm{r}=-0.69, \mathrm{P}<0.0001$ ) at the same time that the North American proportion at Greenland is increasing ( $\mathrm{r}=0.87, \mathrm{P}<0.0001$ ) (Figure 5.3.2). This can only occur if the proportion of southern European salmon migrating to Greenland is declining or if the proportion of North American salmon migrating to Greenland is increasing ( $\mathrm{r}=-0.52, \mathrm{P}<0.004$ ). However, given current trends the former is more likely the case.

### 5.4 Evaluation of the effects on European and North American stocks of the West Greenland management measures since 1993

There have been three significant changes in the management regime at West Greenland since 1993. First, NASCO adopted a new management model (Anon., 1993) based upon ICES' assessment of the PFA of non-maturing 1SW North American salmon and the spawner escapement requirements for these stocks. This resulted in a substantial reduction in the TAC agreed to by NASCO from 840 t in 1991 to 258 t in 1992, and further reductions in subsequent years. The second change in management was the suspension of fishing in 1993 and 1994 following the agreement of compensation payments by the North Atlantic Salmon Fund. Due to the closure of the fishery in the two years no sampling could be carried out in Greenland, and no biological data were collected. The third change in management was a multi-year agreement in 1999 restricting the annual catch to that amount used for internal consumption in Greenland, which in the past has been estimated at 20 t .

To calculate a possible TAC for those years according to the agreed quota allocation model (Anon., 1993) biological parameters from sampling in 1992 were used (Table 5.4.1). The variables given in the table (proportion of origin, mean weights, and proportion of 1SW fish) are those used in the analyses, see Sections 5.1 and 5.7.

The numbers of fish spared by the 1993-1994 closure are shown in Table 5.4.1. The potential catches in the years 1993 and 1994 of 89 and 137 t , respectively correspond to the TACs calculated in accordance with the quota allocation computation model that was agreed by NASCO at its annual meeting in 1993. For the successive years nominal catch figures are used. The table shows the number of salmon returning to home waters provided no fishing of the given magnitude took place in Greenland. The biological parameters given in the table represent the annual sampling data.

The mean number for 1993-2000 of potentially returning fish per ton caught at Greenland is calculated to 208 and 106 salmon for North America and Europe, respectively.

In the years 1972-92 exploitation rates in Greenland of the North American component of the salmon stock fluctuated between 10 and $45 \%$ around an average of $34 \%$ (Figure 5.4.1). The management measures in force since 1993 resulted in an average exploitation rate of this component of $13 \%$, for the period 1995-97, about one third of its previous level after reopening of the fishery in 1995. After the 1999 agreement the exploitation rates decreased to about $5 \%$.

In the current analysis the effects of the management measures taken at West Greenland have been examined in terms of numbers of fish only. Thus it has been difficult to show direct benefits to homewater stocks from these measures. The Working Group recommends that other indices of change, i.e. changes in age composition, size at age and sea survival, should also be included in this evaluation.

### 5.5 Age-Specific Stock Conservation Limits for All Stocks in the West Greenland Commission Area

Sampling of the fishery at West Greenland (Table 5.1.3.3) since 1985 has shown that both European and North American stocks harvested are primarily (greater than $90 \%$ ) 1SW non-maturing salmon that would mature as either 2 or 3 SW salmon, if surviving to spawn. Usually less than $1 \%$ of the harvest are salmon which have previously spawned and a few percent are 2 SW salmon which would mature as 3 SW or older salmon, if surviving to spawn. For example, in $1999,96.8 \%$ of the sampled catch of North American origin and $100 \%$ of the sampled catch of European origin were 1SW salmon. For this reason, conservation limits defined previously for North American stocks have been limited to this cohort (2SW salmon on their return to homewaters) that may have been at Greenland as 1SW non-maturing fish. These numbers have been documented previously by the Working Group and are revised this year in Section 4.4. From Table 4.4.1, the 2SW spawning requirements of salmon stocks from North America which may be present in the West Greenland Commission Area total 152,548 fish, with 123,349 and 29,199 required in Canadian and USA rivers, respectively.

The conservation limits were split into 1SW and MSW components on the basis of the average age composition of catches in the past ten years. The stocks have also been partitioned into northern and southern groups, and tagging information and biological sampling indicates that the majority of the European salmon caught at West Greenland originate from the southern group. The currently estimated conservation limit for southern European MSW stocks is approximately 595,000 fish (Table 3.7.3.1). There is still considerable uncertainty in the conservation limits for European stocks. The above value has been increased from 530,000 in the 2000 report. To date, the conservation limits for MSW salmon in Europe have not been incorporated into the modelling of catch options for West Greenland.

### 5.6.1 Overview of provision of catch advice

The Working Group was asked to advise on catch levels based upon maintaining adequate spawning escapements sufficient to achieve conservation limits. Although advances have been made in our understanding of the population dynamics of Atlantic salmon and the exploitation occurring in the fisheries, the concerns about the implications of applying TACs to mixed-stock fisheries are still relevant. In principle, adjustments in catches in mixed-stock fisheries provided by means of an annually adjusted TAC would reduce mean mortality on the contributing populations. However, benefits that might result for particular stocks would be difficult to demonstrate, in the same way that damage to individual stocks would be difficult to identify.

In 1993, the Working Group considered how the predictive measures of abundance could be used to give annual catch advice (ICES 1993/Assess:10; Sections 5.3 and 5.4). The aim of management would be to limit catches to a level that would facilitate achieving overall spawning escapement reflecting the spawning requirements in individual North American and European rivers (when the latter have been defined). In order to achieve the desired level of exploitation for a given level of predicted abundance, a TAC could be fixed or some form of effort adjustment introduced. Such an assessment would also depend on a forecast of prefishery abundance for both North American and European salmon stocks.

To date, the advice for any given year has been dependent on obtaining a reliable predictor of the abundance of nonmaturing 1SW North American stocks prior to the start of the fishery in Greenland. Gill net fisheries in Greenland and Labrador harvest one-sea-winter (1SW) salmon about one year before they mature and return to spawn in North American rivers. This component was also harvested on their return as 2SW salmon in commercial fisheries in eastern Canada, angling and native fisheries throughout eastern Canada and angling fisheries in the northeastern USA. The fishery in Greenland harvests salmon which would not mature until the following year while the fishery in Labrador (closed in 1998) harvested a mix from the non-maturing component as well as maturing 1SW and MSW salmon. The commercial fisheries in Québec and the Maritime provinces of Canada harvested maturing 1SW and MSW salmon.

The Working Group has advocated models based on thermal habitat in the northwest Atlantic and spawning stock indices to forecast pre-fishery abundance and provide catch advice for the West Greenland fishery. While the approach has been consistent since 1993, the models themselves have varied slightly over the years. The changes have been made to these models in attempts to improve the biological basis and predictive capability. In particular, the models since 1996 have used a spawning stock surrogate variable (lagged spawners) in an attempt to describe the variations in parental stock size of the non-maturing 1SW component (PFA). The models of previous years included the following predictor variables: 1993 - thermal habitat in March; 1994 - thermal habitat in March; 1995 -thermal habitat in January, February, and March; and 1996-2000 - thermal habitat in February and lagged spawners from the Labrador, Newfoundland, Québec, and Scotia-Fundy regions of Canada.

The Working Group noted that because the method of estimating spawning escapement for Labrador was based on commercial catches and exploitation rates ended in 1997, lagged spawner values will have missing components in year 2003. An alternative index of salmon abundance will be required in the future. Preliminary investigations into the development of a juvenile abundance index as an alternative index of salmon abundance were reported in 2000, and continued in the current report (Section 5.8).

## North American run-reconstruction model

The Working Group has used the North American run-reconstruction model to estimate pre-fishery abundance of 1SW non-maturing and maturing 2SW fish adjusted by natural mortality to the time prior to the West Greenland fishery (See Section 4.2.3). Region-specific estimates of 2SW returns are listed in Table 4.2.2.2. Estimates of 2SW returns prior to 1998 in Labrador are derived from estimated 2SW catches in the fishery using a range of assumptions regarding exploitation rates and origin of the catch. With the closure of the Labrador fishery, 1998-2000 returns were estimated as a proportion of the total for other areas based on historical data (Section 4.2.3).

## Update of thermal habitat index

The Working Group has been using the relationship between marine habitat, 2SW lagged spawners and estimated prefishery abundance to forecast pre-fishery abundance in the year of interest (ICES 1993/Assess:10; 1994/Assess:16; 1995/Assess:14; 1996/Assess:11, 1997/Assess:10; 1998/ACFM:15, 1999/ACFM:14; and 2000/ACFM:13). Marine habitat is measured as a relative index of the area suitable for salmon at sea, termed thermal habitat, and was derived from sea surface temperature (SST) data obtained from the National Meteorological Center of the National Ocean \&

Atmospheric Administration and previously published catch rates for salmon from research vessels fishing in the northwest Atlantic (Reddin et al. 1993 and ICES 1995/Assess:14). The SST data were determined by optimally interpolating SSTs from ships of opportunity, earth observation satellites (AVHRR), and sea ice cover data. The area used to determine available salmon habitat encompassed the northwest Atlantic north of $41^{\circ} \mathrm{N}$ latitude and west of $29^{\circ} \mathrm{W}$ longitude and includes the Davis Strait, Labrador Sea, Irminger Sea, and the Grand Bank of Newfoundland.

Thermal habitat index has been updated to include data for 2000 and January and February 2001 year data. Two periods of decline are identified (1980 to 1984 and 1988 to 1995) in the February index (Table 5.6.1.1 and Figure 5.6.1.1). The habitat index for February increased slightly (3\%) in 2001 from 1,634 to 1,685. The 2001 February value is close to the long-term mean of 1,653.

The lagged spawner variable used in the model is an estimate of the 2 SW parental stock of the PFA. The calculation procedure is described in Section 4.2.4. Previous analyses indicated that the sum of lagged spawner components from Labrador, Newfoundland, Québec, and Scotia-Fundy and excluding Gulf and U.S. was the strongest explanatory variable for the model. Inclusion of the Gulf spawning component reduced the explanatory power of the variable.

The Working Group recognized the problems inherent in the exclusion of a major component of the spawning stock contributing to the PFA. As well, spawning escapement estimates for Labrador are not available for the years 19982000. The previously formulated lagged spawner variable will therefore not be available beyond 2002 .

### 5.6.2 Forecast model for pre-fishery abundance of North America 2SW salmon

The model used to forecast pre-fishery abundance for 2001 was revised and results presented in Section 5.6 are based on this model. Changes to the model in 2001 and a comparison in 2001 results produced by each model are provided in Section 5.7. The 2001 forecast of pre-fishery abundance was based on an alternative modelling approach that takes into consideration that habitat acts on PFA through survival rather than on absolute abundance. The adopted model takes the following form:

$$
\text { PFA }=\text { Spawners } * \exp ^{-\left(\alpha+\mathrm{B}^{*} \text { Habitat }+\xi\right)}
$$

This model relates directly to a survival relationship of the form: $N_{t}=N_{0} e^{-Z}$.

In the case of the PFA model, the survival rate of salmon (PFA / Spawners) has a mean survival level which is modified by the habitat environmental variable. A linear form of the model fits the natural $\log$ of PFA relative to the natural $\log$ of spawners and habitat variables:

$$
\operatorname{Ln}(\text { PFA })=\operatorname{Ln}(\text { Spawners })+\text { Habitat }+ \text { intercept }+\xi
$$

The basis for the model is the same two predictor variables as were used from 1999 to 2000: thermal habitat for February (term H2) and lagged spawners (sum of lagged spawners from Labrador, Newfoundland, Scotia-Fundy and Quebec, term SLNQ) (ICES 1996/Assess:11). This was justified on the basis of studies showing that salmon stocks over wide geographic areas tend to have synchronous survival rates and that the winter period may be the critical stage for post-smolt survival and maturation (Scarnecchia 1989; Reddin and Shearer 1987; Friedland et al. 1993; Friedland et al. 1998). Consequently, the input data used in 2000 was updated to reflect the inclusion of the additional value and the refinement of other parameters to the time series of pre-fishery abundance estimates.

There was a significant linear relationship between the estimate PFA values and predicted values (log transformed model; $\mathrm{F}_{2,18}=70.52 ; \mathrm{r}^{2}=0.88$ ). All model parameters were significant at less than the $5 \%$ level (Table 5.6.2.1). Individually, the two predictor variables are also significantly related to pre-fishery abundance. February habitat accounted for approximately $10 \%$ of the total sum of squares and SLNQ spawners was approximately $80 \%$ (Table 5.6.2.1). The jackknife and simulated predicted values for pre-fishery abundance for 1978-2001 are shown in Table 5.6.1.1 and Figure 5.6.2.1. The predicted values are shown to fit the observed data quite well except in the late 1980s and 1990s when abundance was low and there are low positive residuals at the end of the time series (Figure 5.6.2.1). The forecasted estimate by simulation of pre-fishery abundance for 2001 using the February thermal habitat and lagged spawner model is about 295,700 at the $50 \%$ probability level (Table 5.6.1.1).

The model continues to be influenced primarily by the spawning stock level in the predictive relationship for pre-fishery abundance (Table 5.6.2.1). Thus, the prediction of pre-fishery abundance would be moderated during periods of high levels of habitat and low levels of spawning stock. The alternate case would be an increase in predicted pre-fishery abundance when spawning stocks were high and thermal habitat was low. The former has occurred with the predicted
values for 1998 and 1999, as thermal habitat has increased considerably, the predicted pre-fishery abundance in recent years was low due to the large decline in spawners producing them (Figure 5.6.1.1). However, two-sea-winter spawners contributing to returns have improved in the year 2001 which is contributing substantially to the increase in forecasted prefishery abundance.

Using this model to estimate the 2000 pre-fishery abundance yields a value of 225,700 , which is about $25 \%$ higher than the previously reported value of 179,900 . Note that the previously reported value was based on the additive model without errors in the lagged spawners. The inclusion of errors in the lagged spawners has been shown to increase the median value and to widen the distribution of the forecast (ICES CM 2000/ACFM:13). The change is also in part due to the addition of 1999 prefishery abundance which was not included last year as it was unavailable. This value is on the mid- to low end of the distribution of prefishery abundance. Also due to the time lag between forecasted and estimated prefishery abundance there is a delay of two years before comparison of estimated and forecasted values can be made. Consequently, any developing trend in high positive or negative residuals indicating a poor fit to recent data will be hard to detect until after the fishery. It should be noted that deterministic and simulated forecast values will show differences due to the method of calculation.

In Section 4.5.1, the relationship between the available 2 SW to 1 SW data from several rivers in Eastern Canada indicated that the 2000 forecast of prefishery abundances, i.e returns of 2 SW salmon to North America in 2001, is unlikely to be achieved. Consequently, there is considerable uncertainty regarding the projected reversal of the declining trend in prefishery abundance forecasted by the model

## Stochastic Analyses

Although the exact error bounds for the estimates of prefishery abundance (NN1(i)) are unknown, minimum and maximum values of component catch and return estimates have been estimated. Simulation methods, implemented in the software package SAS (SAS Institute, 1996), were used to generate the probability density function of NN1(i). This was done as a seven-step procedure as follows:

Step 1: Annual values (1978-99) of pre-fishery abundance (NN1) were generated assuming a uniform distribution of the minimum to maximum values of input parameters $\mathrm{NC} 1, \mathrm{NC} 2$, and NR 2 .

Step 2: Annual values (1978-99) of lagged spawners (SLNQ) were generated assuming a uniform distribution of the minimum to maximum values of SLNQ

Step 3: The parameter values of the regression model of pre-fishery abundance on the February thermal habitat (H2) variable and the lagged spawners (SLNQ) variable were estimated from the data set generated in steps 1 and 2.

Step 4: A single pre-fishery forecast value for 2001 was obtained by drawing at random from a normal distribution defined by the mean forecast value and the mean square error of the estimate (for a single prediction) from the regression statistics. The normal distribution was used because the error structure of the regression (after log transforamtion) is assumed to be normal.

Step 5: Step 4 was repeated 1,000 times to generate a vector of forecast values from an individual regression fit.

Step 6: Steps 1 to 5 were repeated 1,000 times to generate $1,000,000$ predictions ( 1,000 times 1,000 ) of pre-fishery abundance. This resampling incorporates the uncertainty of the input parameters (step 1 and 2 ) and the unexplained variance in pre-fishery abundance from the regression (step 5).

Step 7: The probability profile of these stochastic realizations (in $5 \%$ intervals) of the pre-fishery abundance forecast was generated from the vector of pre-fishery abundance forecast values obtained in step 6 (Table 5.6.2.2).

These estimates will be used to develop risk analysis and catch advice presented in Section 5.6.3 and 5.6.4. Managers may use this information to determine the relative risks borne by the stock (i.e., not meeting spawning requirements) versus the fishery (e.g., reduced short-term catches).

### 5.6.3 Development of catch options for 2001

Atlantic salmon are managed with the objective of ensuring adequate numbers of spawners in individual rivers. A composite spawning requirement for the North American 2 SW stock complex was developed by summing the spawning
requirements of Salmon Fishing Areas in Canada and river basins within the USA. Details on the methodology to estimate and update the spawning requirements are provided in (ICES 1996/Assess:11) and in Section 4.4 of this report. With these data, it is possible to compute an allowable harvest. This procedure is unchanged from the previous assessment. Previously, NASCO considered all salmon above the conservation requirement as being available for harvest.

The fishery allocation for West Greenland is for 1SW fisheries in 2001, whereas the allocation for North America can be harvested in fisheries on 1SW salmon in 2001 and/or in fisheries on 2SW salmon in 2002. To achieve spawning requirements, a pool of fish must be set aside prior to fishery allocation in order to meet spawning requirements and allow for natural mortality in the intervening months between the fishery and return to river. Thus, 170,286 pre-fishery abundance fish must be reserved $\left(152,548 / \exp ^{\left(-.01^{* 11)}\right)}\right.$ to ensure achievement of the requirement after natural mortality.

Quota computation for the 2001 fishery requires an estimate of pre-fishery abundance [NN1], stock composition by continent [PropNA], mean weights of North American and European 1SW salmon [WT1SWNA and WT1SWE, respectively], and a correction factor for the expected sea-age composition of the total landings [ACF]. Exponentially smoothed values utilising data collected during the 1995-99 fisheries are summarised below.

| Parameter |  |
| :--- | :--- |
|  | Value |
| PropNA | 0.779 |
| WT1SWNA | 2.954 |
| WT1SWE | 2.990 |
| ACF | 1.049 |

Greenland quota options are presented for the $25 \%$ and $75 \%$ cumulative probability levels of PFA (Table 5.6.2.2, Table 5.6.3.1, Appendix 6). The probability distribution provides a measure of the chance that the PFA value would be lower than the value considered. Between the $25 \%$ and $75 \%$ probability level and at the Fna of 0.4 quota options range from 28 to 467 t with a median value of 200 t .

Growth of salmon through the fishing season can significantly affect the total number of fish harvested under a fixed quota. A sensitivity analysis was conducted to evaluate the effect of salmon growth in August and September on the total number of fish harvested under a theoretical 200 mt quota (Figure 5.6.3.1). This analysis shows that the number of fish harvested under a fixed quota declines significantly as the median date of the fishery is delayed through August and September.

### 5.6.4 Risk assessment of catch options for 2001

The provision of catch advice in a risk framework involves the incorporation of the uncertainty in all the factors used to develop the catch options. The method is described in more detail in Section 2.3.1 and 2.3.3. Annual variations in uncertainty result in differing assessments and differing levels of precision. The risk analysis plots are calculated for consideration of the 2001 fishery in West Greenland.

The pre-fishery abundance of salmon in 2001 is predicted to be moderate relative to historic levels (Figure 5.6.4.1). The risk analysis results suggest a moderate risk that the returns of 2SW salmon to North America in 2002 will be below the conservation requirement, even in the absence of any fisheries on this age group in Greenland in 2001 (Figure 5.6.4.2).

The risk analysis performed considers the most optimistic scenario of equal production rates in all six stock areas of North America. The reality is that the stock status differs greatly within North America and that the expected returns of salmon to the USA and Scotia Fundy areas will be severely below their respective conservation requirements of this area. In the USA, the escapement for the entire area has never been above 3000 spawners since 1992, no better than $10 \%$ of the requirement (Table 4.2.4.1). Similarly, the Scotia-Fundy area lagged spawners have been less than 10,000 fish over the last ten years (Table 4.2.4.4). If all stocks were at their spawner requirements, the U.S.A. stocks would be expected to produce almost $20 \%$ of the 2 SW production from North America while the Scotia-Fundy stock is expected to produce just over $16 \%$ of the total (Figure 4.2.4.5). Under the current levels of spawning escapement, recruitment to USA. rivers are not expected to be more than $2 \%$ of the total PFA, and Scotia-Fundy no better than $10 \%$ of the present PFA (Figure 4.2.4.5). The majority of the non-maturing 1SW salmon in the Northwest Atlantic in 2001 are expected to return principally to the other areas, Quebec, Gulf, Labrador and Newfoundland (Figure 4.2.4.5). With this consideration, the risk analysis applies more appropriately to these four areas while the probability of the Scotia-Fundy area meeting its conservation requirement is very likely near zero and is zero for the USA stocks. These differences in anticipated relative production should be considered in the risk analysis for the coming years in an attempt to provide a more realistic evaluation and useful analysis for guiding fisheries management.

The Working Group concludes that the North American stock complex of non-maturing salmon remains in tenuous condition. Increased spawning escapements to rivers of some areas of eastern North America resulted in improved abundance of the juvenile life stages, and perhaps now at adult life stages. Despite the closure of Canadian and West Greenland commercial fisheries, sea survival of adults returning to rivers has not improved and in some areas has declined further. The abundance of maturing 1SW salmon has also declined in many areas of eastern North America. Associations between 1SW returns in year $i$ and 2 SW returns in year $i+1$ observed in several rivers in eastern Canada suggest that abundance of 2SW salmon in 2001 in eastern Canada will be similar to or less than recent years (Sections 4.5.1). Smolt production in 1999 and 2000 in monitored rivers of eastern Canada were similar to or below the average of the last five years and unless sea survival improves, the abundance of non-maturing 1SW salmon in the Northwest Atlantic is not expected to improve above the levels of the last five years.

There is little information available to confirm the possibility of an improvement in prefishery abundance in 2000 and 2001 as forecasted through modelling. Two sea winter adult returns in 2001 will provide initial indications regarding the overall abundance of non-maturing 1 SW salmon in 2000. The adoption of risk neutral quota options on the basis of predicted sharp increases in pre-fishery abundance in 2000 and 2001 provide the potential for significant overexploitation if increases in prefishery abundance are not realized. Extreme caution is urged regarding harvest decisions for 2001, and adoption of conservative harvest levels is warranted until projected increases can be confirmed. The increasing advantage associated with each additional spawner in under-seeded river systems makes a strong case for a conservative management strategy.

### 5.7 Changes to and Critical Assessment of the 'Model' Used to Provide Catch Advice and Impacts of Changes on the Calculated Quota

### 5.7.1 Changes from the 2000 assessment

The models used to predict pre-fishery abundance of the North American non-maturing stock complex and subsequent quota levels for West Greenland were revised based on exploratory work conducted by the Working Group and reported in the 1999 and 2000 reports. For the past several years, models used to predict the PFA hypothesized a linear effect of SLNQ and habitat on salmon abundance.

$$
\text { PFA }=\alpha+\beta^{*} \text { SLNQ }+\gamma^{*} \text { Habitat }+\xi
$$

An alternative approach, adopted in the 2001 assessment takes into consideration that habitat acts on PFA by mediating survival rather than on absolute abundance. The adopted model takes the following form:

$$
\text { PFA }=\quad \beta * \text { SLNQ } * \exp ^{-(\alpha+\gamma * \text { Habitat }+\xi)}
$$

This model relates directly to a survival relationship of the form:

$$
\mathrm{N}_{\mathrm{t}}=\mathrm{N}_{0} \mathrm{e}^{-\mathrm{Z}}
$$

where

- $\quad \mathrm{N}_{\mathrm{t}}$ is the abundance at time t (PFA)
- $\mathrm{N}_{0}$ is the abundance at time 0 (SLNQ)
- Z is the instantaneous mortality rate (at a mean level with an environmental modification)

A linear form of the model fits the natural $\log$ of PFA relative to the natural $\log$ of spawners and habitat variables:

$$
\ln (\mathrm{PFA})=\alpha+\beta^{\prime} * \operatorname{Ln}(\mathrm{SLNQ})+\gamma^{*} \text { Habitat }+\xi
$$

where

- $\xi$ is assumed $\mathrm{N}(0, \sigma)$

Note that under the back-transformed model, the error structure is assumed lognormal.

The basis for the multiplicative model is two predictor variables: thermal habitat for February (term H2) and lagged spawners (sum of lagged spawners from Labrador, Newfoundland, Scotia-Fundy and Quebec, term SLNQ; ICES 1996/Assess:11), which are unchanged from the 2000 assessment.

In addition, the uncertainty in the lagged spawner and PFA variables were incorporated in the model simulations. A triangular distribution centered at the midpoints and defined by the minimum and maximum values was assumed for the uncertainty in the PFA and lagged spawner variables. Values were drawn randomly and independently for each variable, each year in the simulations. A total of 1,000 simulations were performed to generate the posterior predictive probability distributions of the PFA in year 2000. When predicting the PFA, the uncertainty in the lagged spawners for the year 2000 was also considered in the same way as for the predictor variables.

There is an important difference in distribution between the additive model used in the 2000 assessment and multiplicative models (adopted in 2001) with the latter being skewed to the origin and long tailed towards large values. Because of the underlying lognormal distribution of errors, the predicted abundance is always greater than 0 , contrary to what is given by the additive model. There is a greater cumulative probability for lower PFA levels with the multiplicative model but the distribution suggests that there is insufficient information in the data to fix an upper bound on the PFA. Adoption of the multiplicative model resolves issues related to the biological logic of the model and the prediction of unreasonable PFA values, less than 0 , generated by the additive model in previous assessments.

The multiplicative model has improved statistical fit to the two predictor variables used in previous assessments. Figure 5.7.1.1 provides a comparison of the residuals and pre-fishery abundance estimates for the additive model used in the 2000 assessment and the multiplicative model adopted in 2001. Each model provides similar estimates of pre-fishery abundance and associated residuals. The primary changes in performance of the multiplicative model adopted for 2001 relate to changes in pre-fishery abundance forecasts and characterization of uncertainty about these forecasts.

### 5.7.2 Impact of changes on the catch advice

Adopting the multiplicative model does result in changes to the catch advice generated. Table 5.7.2.1 provides the probability levels of the pre-fishery abundance estimates for the additive model used in the 2000 assessment. The additive model provides 2001 forecasts of 312,000 and 425,000 salmon at $25 \%$ and $75 \%$ probability levels, respectively. The multiplicative model generates prefishery abundance forecasts for 2001 of 187,700 and 463,000 salmon at the $25 \%$ and $75 \%$ probability levels, respectively (Table 5.6.2.2). The distribution of PFA forecasts generated by the multiplicative model is centered at a lower median and has a broader distribution of probability values, particularly in the end of PFA.

Table 5.7.2.2 provides quota options for 2001 at West Greenland based on results from the pre-fishery abundance estimates generated by the additive model used in the 2000 assessment. The additive model used in the 2000 assessment provides a pre-fishery abundance forecast for non-maturing 1SW salmon in 2001 of 368,685 , while the multiplicative model adopted in 2001 provides a lower forecast of 295,678 (approximately $20 \%$ lower). In comparing quota options generated by the multiplicative and additive models for a $40 \%$ portion of the surplus fishery abundance to West Greenland at a $50 \%$ probability level, the multiplicative model provides a quota option of 200 t (Table 5.6.3.1), while the additive model provides a quota option of 317 t (Table 5.7.2.2).

## Alternative Spawning Stock Variable

As an alternative to the lagged spawner variable, juvenile abundance indices were considered as surrogates of potential smolt production from eastern Canada as described in Section 4.2.1. The adjustment (annual value divided by the average value of 1994 to 1998) places all the rivers on a common relative scale, temporally. The information from various rivers with juvenile surveys or smolt counts is combined into an index of smolt production reflecting the temporal variability in the juvenile production (Section 4.2.1).

The index of smolts from North America was obtained by weighting the annual river indices by the relative proportion of the conservation egg requirements (O'Connell et al. 1997) of the SFA or Zone to the total conservation egg requirements of the zones under consideration (Table 4.2.1.2). An alternative weighting incorporated the relative contribution to the 2SW spawner requirements of the six main areas within North America. This allows indices of smolt production from all areas of North America to be used but attributes weights to the area indices according to the expected contribution to 2 SW abundance.

The relative indices using alternative weighting factors all show the same general pattern of freshwater production: at about one-third the recent (1995 to 1998) average between 1971 and 1979, at about $60 \%$ of the average during 1980 to 1985 and at about average since 1986 (Figure 4.2.1.6).

A model identical to that described in sections 5.6 and 5.7 was fitted after substituting the juvenile index for the lagged spawner variable (SLNQ). The juvenile index was advanced one year to correspond to the year of PFA (i.e. the PFA of year i corresponded to the juvenile index of year i-1 which was a combination of smolt indices of year i-1 and the parr indices of year i-2).

The modeled relationship between juvenile index and PFA is negative indicating that as the juvenile index increases, PFA decreases (Figure 5.8.1). There has not been any temporal contrast in juvenile index and PFA values such that a generally increasing trend in juveniles corresponds to the generally decreasing trend in PFA over the time series examined.

In terms of the predictions for the PFA, the 2000 value has a median estimate of 120,000 fish, $50 \%$ lower than the value derived from the lagged spawner model without errors in spawners (Figure 5.8.2) (ICES CM 2000/ACFM:13). For 2001, the model provides the following predictions of PFA; median $=122,000$ fish with a $10^{\text {th }}$ to $90^{\text {th }}$ percentile range of 71,000 to 216,000 fish (Figure 5.8.2).

### 5.8 Concerns regarding the juvenile index

A juvenile index model is conceptually more attractive as juveniles represent a stage closer to the PFA than the lagged spawner variable used previously. Consequently, some of the noise corresponding to the stochasticity of the recruitment process between the spawner and the juveniles is removed, favoring a more direct link between the predictors and the PFA.

This point needs to be addressed because in many cases (especially until the mid 1980s) the index is based on only a few rivers, and the rivers monitored never represented more than $35 \%$ of the total area potentially producing juveniles. The number of sampling stations by river is also limited, whereas juvenile abundance is know to show small-scale spatial variations and the measurement made on each station is an estimate with an uncertainty associated with it. As a first step, a sensitivity analysis of the PFS forecast to measurement errors in juvenile indices is required before applying the index in a predictive framework for PFA abundance.

The rivers monitored for juveniles are assumed to represent the relative production levels within a broader geographic area. This assumption should be tested by examining trends in abundance where several rivers are available from a zone.

The juvenile index also assumes that parr to smolt translations are equivalent in all areas. This should also be examined where parr and smolt data sets from the same river are available. Parr size-at-age has been shown to vary annually in the Miramichi River (Section 2.4.7) and size has been shown to be an important determinant of smoltification. Consequently, increased juvenile densities may not translate directly into smolts especially where overwinter survival of large parr has been shown to be in some rivers limiting smolt production

As with the other indices of spawning stock, there is an assumption of stationarity over time in parr to smolt dynamics. Again, where data sets exist, this should be examined.

### 5.8.1 Alternative modelling approaches

All the models examined to date assume that the habitat, spawning stock indicators and PFA estimates are temporally independent. In reality, all these data sets are time series with autocorrelation (as evidenced in residual patterns). Models to treat time series data should be examined.

There is also the potential problem of non-stationarity in the data sets being examined. Examples from both sides of the Atlantic provide evidence of shifts in marine survival over the few decades of observations available. Models such as dynamic linear modeling would permit the integration of this information sequentially through time. It would be useful for the Working Group to review these approaches in the near future to address the various problems identified with the modeling approaches to date.

### 5.9 Data Deficiencies and Research Needs in the WGC area

1. Continued efforts should be made to improve the estimates of the annual catches of salmon taken for local consumption in Greenland.
2. The mean weights, sea and freshwater ages and continent of origin are essential parameters to provide catch advice for the West Greenland fishery. As these parameters are known to vary over time, the Working Group recommends that the sampling programme be continued and closely coordinated with fishery harvest plan to be executed annually in West Greenland.
3. The catch options for the West Greenland fishery are based almost entirely upon data taken from North American stocks (with the current exclusion of Labrador, see Section 4.6). In view of the evidence of a long-term decline in the European stock components contributing to this fishery (southern European non-maturing 1SW recruits) the Working Group emphasised the need for information from these stocks to be incorporated into the assessments as soon as possible.
4. Alternative models should be explored (for example different predictive variables, model formulations, univariate time series, non-parametric change-of-state analyses) to provide some index of plausibility of the quantitative forecasts.
5. Further basic research is needed on the spatial/temporal distribution and migration patterns of salmon and their predators at sea to assist in explaining variability in survival rates.
6. Samples should be obtained for DNA analysis from rivers in North America and Europe.

The status of the six stock areas should be incorporated into the analysis of risk of catch options.

Table 5.1.1. Nominal catches of salmon, West Greenland 1960-2000 (metric tons round fresh weight).

| Year | Norway | Faroes | Sweden | Denmark | Greenland ${ }^{1}$ | Total | Quota ${ }^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | - | - | - | - | 60 | 60 | - |
| 1961 | - | - | - | - | 127 | 127 | - |
| 1962 | - | - | - | - | 244 | 244 | - |
| 1963 | - | - | - | - | 466 | 466 | - |
| 1964 | - | - | - | - | 1539 | 1539 | - |
| 1965 | $-3^{3}$ | 36 | - | - | 825 | 861 | - |
| 1966 | 32 | 87 | - | - | 1251 | 1370 | - |
| 1967 | 78 | 155 | - | 85 | 1283 | 1601 | - |
| 1968 | 138 | 134 | 4 | 272 | 579 | 1127 | - |
| 1969 | 250 | 215 | 30 | 355 | 1360 | 2210 | - |
| 1970 | 270 | 259 | 8 | 358 | 1244 | $2146{ }^{4}$ | - |
| 1971 | 340 | 255 | - | 645 | 1449 | 2689 | - |
| 1972 | 158 | 144 | - | 401 | 1410 | 2113 | 1100 |
| 1973 | 200 | 171 | - | 385 | 1585 | 2341 | 1100 |
| 1974 | 140 | 110 | - | 505 | 1162 | 1917 | 1191 |
| 1975 | 217 | 260 | - | 382 | 1171 | 2030 | 1191 |
| 1976 | - | - | - | - | 1175 | 1175 | 1191 |
| 1977 | - | - | - | - | 1420 | 1420 | 1191 |
| 1978 | - | - | - | - | 984 | 984 | 1191 |
| 1979 | - | - | - | - | 1395 | 1395 | 1191 |
| 1980 | - | - | - | - | 1194 | 1194 | 1191 |
| 1981 | - | - | - | - | 1264 | 1264 | $1265{ }^{6}$ |
| 1982 | - | - | - | - | 1077 | 1077 | $1253{ }^{6}$ |
| 1983 | - | - | - | - | 310 | 310 | 1191 |
| 1984 | - | - | - | - | 297 | 297 | 870 |
| 1985 | - | - | - | - | 864 | 864 | 852 |
| 1986 | - | - | - | - | 960 | 960 | 909 |
| 1987 | - | - | - | - | 966 | 966 | 935 |
| 1988 | - | - | - | - | 893 | 893 | -7 |
| 1989 | - | - | - | - | 337 | 337 | 7 |
| 1990 | - | - | - | - | 274 | 274 | -7 |
| 1991 | - | - | - | - | 472 | 472 | 840 |
| 1992 | - | - | - | - | 237 | 237 | $258{ }^{8}$ |
| 1993 | - | - | - | - | $0^{5}$ | $0{ }^{5}$ | $89^{9}$ |
| 1994 | - | - | - | - | $0^{5}$ | $0{ }^{5}$ | $137{ }^{9}$ |
| 1995 | - | - | - | - | 83 | 83 | 77 |
| 1996 | - | - | - | - | 92 | 92 | $174{ }^{8}$ |
| 1997 | - | - | - | - | 58 | 58 | 57 |
| 1998 | - | - | - | - | 11 | 11 | $20^{10}$ |
| 1999 | - | - | - | - | 19 | 19 | $20^{10}$ |
| 2000 | - | - | - | - | 21 | 21 | $20^{10}$ |

[^11]Table 5.1.1.2. Distribution of nominal catches (metric tons), Greenland vessels.

${ }^{1}$ ) The fishery was suspended
+) Small catches $<0.5$ t
-) No commercial landings

Table 5.1.2.1. Size of biological samples and percentage (by number) of North American and European salmon in research vessel catches at West Greenland (1969-82), from commercial samples (1978-92 and 1995-97), and from local consumption samples (1998-2000).

| Source | Year | Sample size |  | Continent of origin (\%) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Length | Scales | NA | $(95 \% \mathrm{CI})^{1}$ | E | $(95 \% \mathrm{CI})^{1}$ |
| Research | 1969 | 212 | 212 | 51 | $(57,44)$ | 49 | $(56,43)$ |
|  | 1970 | 127 | 127 | 35 | $(43,26)$ | 65 | $(75,57)$ |
|  | 1971 | 247 | 247 | 34 | $(40,28)$ | 66 | $(72,50)$ |
|  | 1972 | 3488 | 3488 | 36 | $(37,34)$ | 64 | $(66,63)$ |
|  | 1973 | 102 | 102 | 49 | $(59,39)$ | 51 | $(61,41)$ |
|  | 1974 | 834 | 834 | 43 | $(46,39)$ | 57 | $(61,54)$ |
|  | 1975 | 528 | 528 | 44 | $(48,40)$ | 56 | $(60,52)$ |
|  | 1976 | 420 | 420 | 43 | $(48,38)$ | 57 | $(62,52)$ |
|  | 1977 | - | - | 45 | ( | 55 | ( |
|  | $1978{ }^{2}$ | 606 | 606 | 38 | $(41,34)$ | 62 | $(66,59)$ |
|  | $1978{ }^{3}$ | 49 | 49 | 55 | $(69,41)$ | 45 | $(59,31)$ |
|  | 1979 | 328 | 328 | 47 | $(52,41)$ | 53 | $(59,48)$ |
|  | 1980 | 617 | 617 | 58 | $(62,54)$ | 42 | $(46,38)$ |
|  | 1982 | 443 | 443 | 47 | $(52,43)$ | 53 | $(58,48)$ |
| Commerc ial | 1978 | 392 | 392 | 52 | $(57,47)$ | 48 | $(53,43)$ |
|  | 1979 | 1653 | 1653 | 50 | $(52,48)$ | 50 | $(52,48)$ |
|  | 1980 | 978 | 978 | 48 | $(51,45)$ | 52 | $(55,49)$ |
|  | 1981 | 4570 | 1930 | 59 | $(61,58)$ | 41 | $(42,39)$ |
|  | 1982 | 1949 | 414 | 62 | $(64,60)$ | 38 | $(40,36)$ |
|  | 1983 | 4896 | 1815 | 40 | $(41,38)$ | 60 | $(62,59)$ |
|  | 1984 | 7282 | 2720 | 50 | $(53,47)$ | 50 | $(53,47)$ |
|  | 1985 | 13272 | 2917 | 50 | $(53,46)$ | 50 | $(54,47)$ |
|  | 1986 | 20394 | 3509 | 57 | $(66,48)$ | 43 | $(52,34)$ |
|  | 1987 | 13425 | 2960 | 59 | $(63,54)$ | 41 | $(46,37)$ |
|  | 1988 | 11047 | 2562 | 43 | $(49,38)$ | 57 | $(62,51)$ |
|  | 1989 | 9366 | 2227 | 56 | $(60,52)$ | 44 | $(48,40)$ |
|  | 1990 | 4897 | 1208 | 75 | $(79,70)$ | 25 | $(30,21)$ |
|  | 1991 | 5005 | 1347 | 65 | $(69,61)$ | 35 | $(39,31)$ |
|  | 1992 | 6348 | 1648 | 54 | $(57,50)$ | 46 | $(50,43)$ |
|  | 1995 | 2045 | 2045 | 68 | $(72,65)$ | 32 | $(35,28)$ |
|  | 1996 | 3341 | 1297 | 73 | $(76,71)$ | 27 | $(29,24)$ |
|  | 1997 | 794 | 282 | 80 | $(84,75)$ | 20 | $(25,16)$ |
| Local cons. | 1998 | 540 | 406 | 79 | $(84,73)$ | 21 | $(27,16)$ |
|  | 1999 | 532 | 532 | 90 | $(97,84)$ | 10 | $(16,3)$ |
|  | 2000 | 491 | 491 | 70 |  | 30 |  |

[^12]Table 5.1.2.2. The weighted proportions and numbers of North American and European Atlantic salmon caught at West Greenland 1982-1992 and 1995-2000. Numbers are rounded to the nearest hundred fish.

|  | Proportion weighted <br> by catch in number |  | E |  |
| :---: | :---: | ---: | ---: | ---: |
| Year | NA |  |  | Numbers of Salmon caught |

Table 5.1.3.1. Annual mean fork lengths and whole weights of Atlantic salmon caught at West Greenland, 1969-1992 and 1995-2000.
Fork length (cm); whole weight (kg). NA = North America; E = Europe.

| Year | Whole weight (kg) |  |  |  |  |  |  |  |  | Fork length (cm) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sea age \& origin |  |  |  |  |  |  |  |  | Sea age \& origin |  |  |  |  |  |
|  | 1SW |  | 2SW |  | PS |  | All sea ages |  | TOTAL | 1SW |  | 2SW |  | PS |  |
|  | NA | E | NA | E | NA | E | NA | E |  | NA | E | NA | E | NA | E |
| 1969 | 3.12 | 3.76 | 5.48 | 5.80 | - | 5.13 | 3.25 | 3.86 | 3.58 | 65.0 | 68.7 | 77.0 | 80.3 | - | 75.3 |
| 1970 | 2.85 | 3.46 | 5.65 | 5.50 | 4.85 | 3.80 | 3.06 | 3.53 | 3.28 | 64.7 | 68.6 | 81.5 | 82.0 | 78.0 | 75.0 |
| 1971 | 2.65 | 3.38 | 4.30 | - | - | - | 2.68 | 3.38 | 3.14 | 62.8 | 67.7 | 72.0 | - | - | - |
| 1972 | 2.96 | 3.46 | 5.85 | 6.13 | 2.65 | 4.00 | 3.25 | 3.55 | 3.44 | 64.2 | 67.9 | 80.7 | 82.4 | 61.5 | 69.0 |
| 1973 | 3.28 | 4.54 | 9.47 | 10.00 | - | - | 3.83 | 4.66 | 4.18 | 64.5 | 70.4 | 88.0 | 96.0 | 61.5 | - |
| 1974 | 3.12 | 3.81 | 7.06 | 8.06 | 3.42 | - | 3.22 | 3.86 | 3.58 | 64.1 | 68.1 | 82.8 | 87.4 | 66.0 | - |
| 1975 | 2.58 | 3.42 | 6.12 | 6.23 | 2.60 | 4.80 | 2.65 | 3.48 | 3.12 | 61.7 | 67.5 | 80.6 | 82.2 | 66.0 | 75.0 |
| 1976 | 2.55 | 3.21 | 6.16 | 7.20 | 3.55 | 3.57 | 2.75 | 3.24 | 3.04 | 61.3 | 65.9 | 80.7 | 87.5 | 72.0 | 70.7 |
| 1977 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 1978 | 2.96 | 3.50 | 7.00 | 7.90 | 2.45 | 6.60 | 3.04 | 3.53 | 3.35 | 63.7 | 67.3 | 83.6 | - | 60.8 | 85.0 |
| 1979 | 2.98 | 3.50 | 7.06 | 7.60 | 3.92 | 6.33 | 3.12 | 3.56 | 3.34 | 63.4 | 66.7 | 81.6 | 85.3 | 61.9 | 82.0 |
| 1980 | 2.98 | 3.33 | 6.82 | 6.73 | 3.55 | 3.90 | 3.07 | 3.38 | 3.22 | 64.0 | 66.3 | 82.9 | 83.0 | 67.0 | 70.9 |
| 1981 | 2.77 | 3.48 | 6.93 | 7.42 | 4.12 | 3.65 | 2.89 | 3.58 | 3.17 | 62.3 | 66.7 | 82.8 | 84.5 | 72.5 | - |
| 1982 | 2.79 | 3.21 | 5.59 | 5.59 | 3.96 | 5.66 | 2.92 | 3.43 | 3.11 | 62.7 | 66.2 | 78.4 | 77.8 | 71.4 | 80.9 |
| 1983 | 2.54 | 3.01 | 5.79 | 5.86 | 3.37 | 3.55 | 3.02 | 3.14 | 3.10 | 61.5 | 65.4 | 81.1 | 81.5 | 68.2 | 70.5 |
| 1984 | 2.64 | 2.84 | 5.84 | 5.77 | 3.62 | 5.78 | 3.20 | 3.03 | 3.11 | 62.3 | 63.9 | 80.7 | 80.0 | 69.8 | 79.5 |
| 1985 | 2.50 | 2.89 | 5.42 | 5.45 | 5.20 | 4.97 | 2.72 | 3.01 | 2.87 | 61.2 | 64.3 | 78.9 | 78.6 | 79.1 | 77.0 |
| 1986 | 2.75 | 3.13 | 6.44 | 6.08 | 3.32 | 4.37 | 2.89 | 3.19 | 3.03 | 62.8 | 65.1 | 80.7 | 79.8 | 66.5 | 73.4 |
| 1987 | 3.00 | 3.20 | 6.36 | 5.96 | 4.69 | 4.70 | 3.10 | 3.26 | 3.16 | 64.2 | 65.6 | 81.2 | 79.6 | 74.8 | 74.8 |
| 1988 | 2.83 | 3.36 | 6.77 | 6.78 | 4.75 | 4.64 | 2.93 | 3.41 | 3.18 | 63.0 | 66.6 | 82.1 | 82.4 | 74.7 | 73.8 |
| 1989 | 2.56 | 2.86 | 5.87 | 5.77 | 4.23 | 5.83 | 2.77 | 2.99 | 2.87 | 62.3 | 64.5 | 80.8 | 81.0 | 73.8 | 82.2 |
| 1990 | 2.53 | 2.61 | 6.47 | 5.78 | 3.90 | 5.09 | 2.67 | 2.72 | 2.69 | 62.3 | 62.7 | 83.4 | 81.1 | 72.6 | 78.6 |
| 1991 | 2.42 | 2.54 | 5.82 | 6.23 | 5.15 | 5.09 | 2.57 | 2.79 | 2.65 | 61.6 | 62.7 | 80.6 | 82.2 | 81.7 | 80.0 |
| 1992 | 2.54 | 2.66 | 6.49 | 6.01 | 4.09 | 5.28 | 2.86 | 2.74 | 2.81 | 62.3 | 63.2 | 83.4 | 81.1 | 77.4 | 82.7 |
| 1995 | 2.37 | 2.67 | 6.09 | 5.88 | 3.71 | 4.98 | 2.45 | 2.75 | 2.56 | 61.0 | 63.2 | 81.3 | 81.0 | 70.9 | 81.3 |
| 1996 | 2.63 | 2.86 | 6.50 | 6.30 | 4.98 | 5.44 | 2.83 | 2.90 | 2.88 | 62.8 | 64.0 | 81.4 | 81.1 | 77.1 | 79.4 |
| 1997 | 2.57 | 2.82 | 7.95 | 6.11 | 4.82 | 6.90 | 2.63 | 2.84 | 2.71 | 62.3 | 63.6 | 85.7 | 84.0 | 79.4 | 87.0 |
| 1998 | 2.72 | 2.83 | 6.44 | - | 3.28 | 4.77 | 2.76 | 2.84 | 2.78 | 62.0 | 62.7 | 84.0 | - | 66.3 | 76.0 |
| 1999 | 3.02 | 3.03 | 7.59 | - | 4.20 | - | 3.09 | 3.03 | 3.08 | 63.8 | 63.5 | 86.6 | - | 70.9 | - |
| 2000 | 2.47 | 2.81 | - | - | 2.58 | - | 2.47 | 2.81 | 2.57 | 60.7 | 63.2 | - | - | 64.7 | - |

Table 5.1.3.2. River age distribution (\%) for all North American and European origin salmon caught at West Greenland, 1968-1992 and 1995-2000.

| River age |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| North American |  |  |  |  |  |  |  |  |
| 1968 | 0.3 | 19.6 | 40.4 | 21.3 | 16.2 | 2.2 | 0.0 | 0.0 |
| 1969 | 0.0 | 27.1 | 45.8 | 19.6 | 6.5 | 0.9 | 0.0 | 0.0 |
| 1970 | 0.0 | 58.1 | 25.6 | 11.6 | 2.3 | 2.3 | 0.0 | 0.0 |
| 1971 | 1.2 | 32.9 | 36.5 | 16.5 | 9.4 | 3.5 | 0.0 | 0.0 |
| 1972 | 0.8 | 31.9 | 51.4 | 10.6 | 3.9 | 1.2 | 0.4 | 0.0 |
| 1973 | 2.0 | 40.8 | 34.7 | 18.4 | 2.0 | 2.0 | 0.0 | 0.0 |
| 1974 | 0.9 | 36.0 | 36.6 | 12.0 | 11.7 | 2.6 | 0.3 | 0.0 |
| 1975 | 0.4 | 17.3 | 47.6 | 24.4 | 6.2 | 4.0 | 0.0 | 0.0 |
| 1976 | 0.7 | 42.6 | 30.6 | 14.6 | 10.9 | 0.4 | 0.4 | 0.0 |
| 1977 | - | - | - | - | - | - | - | - |
| 1978 | 2.7 | 31.9 | 43.0 | 13.6 | 6.0 | 2.0 | 0.9 | 0.0 |
| 1979 | 4.2 | 39.9 | 40.6 | 11.3 | 2.8 | 1.1 | 0.1 | 0.0 |
| 1980 | 5.9 | 36.3 | 32.9 | 16.3 | 7.9 | 0.7 | 0.1 | 0.0 |
| 1981 | 3.5 | 31.6 | 37.5 | 19.0 | 6.6 | 1.6 | 0.2 | 0.0 |
| 1982 | 1.4 | 37.7 | 38.3 | 15.9 | 5.8 | 0.7 | 0.0 | 0.2 |
| 1983 | 3.1 | 47.0 | 32.6 | 12.7 | 3.7 | 0.8 | 0.1 | 0.0 |
| 1984 | 4.8 | 51.7 | 28.9 | 9.0 | 4.6 | 0.9 | 0.2 | 0.0 |
| 1985 | 5.1 | 41.0 | 35.7 | 12.1 | 4.9 | 1.1 | 0.1 | 0.0 |
| 1986 | 2.0 | 39.9 | 33.4 | 20.0 | 4.0 | 0.7 | 0.0 | 0.0 |
| 1987 | 3.9 | 41.4 | 31.8 | 16.7 | 5.8 | 0.4 | 0.0 | 0.0 |
| 1988 | 5.2 | 31.3 | 30.8 | 20.9 | 10.7 | 1.0 | 0.1 | 0.0 |
| 1989 | 7.9 | 39.0 | 30.1 | 15.9 | 5.9 | 1.3 | 0.0 | 0.0 |
| 1990 | 8.8 | 45.3 | 30.7 | 12.1 | 2.4 | 0.5 | 0.1 | 0.0 |
| 1991 | 5.2 | 33.6 | 43.5 | 12.8 | 3.9 | 0.8 | 0.3 | 0.0 |
| 1992 | 6.7 | 36.7 | 34.1 | 19.1 | 3.2 | 0.3 | 0.0 | 0.0 |
| 1995 | 2.4 | 19.0 | 45.4 | 22.6 | 8.8 | 1.8 | 0.1 | 0.0 |
| 1996 | 1.7 | 18.7 | 46.0 | 23.8 | 8.8 | 0.8 | 0.1 | 0.0 |
| 1997 | 1.3 | 16.4 | 48.4 | 17.6 | 15.1 | 1.3 | 0.0 | 0.0 |
| 1998 | 4.0 | 35.1 | 37.0 | 16.5 | 6.1 | 1.1 | 0.1 | 0.0 |
| 1999 | 2.7 | 23.5 | 50.6 | 20.3 | 2.9 | 0.0 | 0.0 | 0.0 |
| 2000 | 3.2 | 26.6 | 38.6 | 23.4 | 7.6 | 0.6 | 0.0 | 0.0 |
| Mean | 3.9 | 34.8 | 37.3 | 16.7 | 6.1 | 1.1 | 0.1 | 0.0 |

Table 5.1.3.2. Continued
European River age

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1968 | 21.6 | 60.3 | 15.2 | 2.7 | 0.3 | 0.0 | 0.0 | 0.0 |
| 1969 | 0.0 | 83.8 | 16.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1970 | 0.0 | 90.4 | 9.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1971 | 9.3 | 66.5 | 19.9 | 3.1 | 1.2 | 0.0 | 0.0 | 0.0 |
| 1972 | 11.0 | 71.2 | 16.7 | 1.0 | 0.1 | 0.0 | 0.0 | 0.0 |
| 1973 | 26.0 | 58.0 | 14.0 | 2.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1974 | 22.9 | 68.2 | 8.5 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1975 | 26.0 | 53.4 | 18.2 | 2.5 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1976 | 23.5 | 67.2 | 8.4 | 0.6 | 0.3 | 0.0 | 0.0 | 0.0 |
| 1977 | - | - | - | - | - | - | - | - |
| 1978 | 26.2 | 65.4 | 8.2 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1979 | 23.6 | 64.8 | 11.0 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1980 | 25.8 | 56.9 | 14.7 | 2.5 | 0.2 | 0.0 | 0.0 | 0.0 |
| 1981 | 15.4 | 67.3 | 15.7 | 1.6 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1982 | 15.6 | 56.1 | 23.5 | 4.2 | 0.7 | 0.0 | 0.0 | 0.0 |
| 1983 | 34.7 | 50.2 | 12.3 | 2.4 | 0.3 | 0.1 | 0.1 | 0.0 |
| 1984 | 22.7 | 56.9 | 15.2 | 4.2 | 0.9 | 0.2 | 0.0 | 0.0 |
| 1985 | 20.2 | 61.6 | 14.9 | 2.7 | 0.6 | 0.0 | 0.0 | 0.0 |
| 1986 | 19.5 | 62.5 | 15.1 | 2.7 | 0.2 | 0.0 | 0.0 | 0.0 |
| 1987 | 19.2 | 62.5 | 14.8 | 3.3 | 0.3 | 0.0 | 0.0 | 0.0 |
| 1988 | 18.4 | 61.6 | 17.3 | 2.3 | 0.5 | 0.0 | 0.0 | 0.0 |
| 1989 | 18.0 | 61.7 | 17.4 | 2.7 | 0.3 | 0.0 | 0.0 | 0.0 |
| 1990 | 15.9 | 56.3 | 23.0 | 4.4 | 0.2 | 0.2 | 0.0 | 0.0 |
| 1991 | 20.9 | 47.4 | 26.3 | 4.2 | 1.2 | 0.0 | 0.0 | 0.0 |
| 1992 | 11.8 | 38.2 | 42.8 | 6.5 | 0.6 | 0.0 | 0.0 | 0.0 |
| 1995 | 14.8 | 67.3 | 17.2 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1996 | 15.8 | 71.1 | 12.2 | 0.9 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1997 | 4.1 | 58.1 | 37.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 1998 | 28.6 | 60.0 | 7.6 | 2.9 | 0.0 | 1.0 | 0.0 | 0.0 |
| 1999 | 27.7 | 65.1 | 7.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2000 | 36.5 | 46.7 | 13.1 | 2.9 | 0.7 | 0.0 | 0.0 | 0.0 |
| Mean | 20.2 | 60.9 | 16.2 | 2.4 | 0.3 | 0.0 | 0.0 | 0.0 |

Table 5.1.3.3. Sea-age composition (\%) of samples from commercial catches at West Greenland, 1985-2000.

|  | North American |  |  | European |  |  |
| :---: | ---: | ---: | :---: | ---: | ---: | ---: |
| Year | Srevious |  | Previous <br> spawners |  |  |  |
| 1985 | 92.5 | 7.2 | 0.3 | 95.0 | 4.7 | 0.4 |
| 1986 | 95.1 | 3.9 | 1.0 | 97.5 | 1.9 | 0.6 |
| 1987 | 96.3 | 2.3 | 1.4 | 98.0 | 1.7 | 0.3 |
| 1988 | 96.7 | 2.0 | 1.2 | 98.1 | 1.3 | 0.5 |
| 1989 | 92.3 | 5.2 | 2.4 | 95.5 | 3.8 | 0.6 |
| 1990 | 95.7 | 3.4 | 0.9 | 96.3 | 3.0 | 0.7 |
| 1991 | 95.6 | 4.1 | 0.4 | 93.4 | 6.5 | 0.2 |
| 1992 | 91.9 | 8.0 | 0.1 | 97.5 | 2.1 | 0.4 |
| 1993 | - | - | - | - | - | - |
| 1994 | - | - | - | - | - | - |
| 1995 | 96.8 | 1.5 | 1.7 | 97.3 | 2.2 | 0.5 |
| 1996 | 94.1 | 3.8 | 2.1 | 96.1 | 2.7 | 1.2 |
| 1997 | 98.2 | 0.6 | 1.2 | 99.3 | 0.4 | 0.4 |
| $1998^{1}$ | 96.8 | 0.5 | 2.7 | 99.4 | 0.0 | 0.6 |
| $1999^{1}$ | 96.8 | 1.2 | 2.0 | 100.0 | 0.0 | 0.0 |
| $2000^{1}$ | 97.4 | 0.0 | 2.6 | 100.0 | 0.0 | 0.0 |

${ }^{1}$ Catches for local consumption only.
Table 5.4.1. Numbers of salmon returning to home waters provided no fishing took place at Greenland. The average number of potentially returning salmon per ton caught in Greenland is also given.

| Year | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Nominal catch at Greenland (tons) ${ }^{1}:$ | 89 | 137 | 83 | 92 | 58 | 11 | 19 | 21 |
| Proportion of NA fish in catch (PropNA): | 0.540 | 0.540 | 0.680 | 0.732 | 0.796 | 0.785 | 0.900 | 0.700 |
| Proportion of EU fish in catch (PropEU): | 0.460 | 0.460 | 0.320 | 0.268 | 0.204 | 0.215 | 0.100 | 0.300 |
| Mean weight, NA fish, all sea ages (kg): | 2.655 | 2.655 | 2.450 | 2.830 | 2.630 | 2.760 | 3.090 | 3.020 |
| Mean weight, EU fish, all sea ages (kg): | 2.745 | 2.745 | 2.750 | 2.900 | 2.840 | 2.840 | 3.030 | 2.992 |
| Mean weight of all sea ages (NA+EU fish): | 2.696 | 2.696 | 2.546 | 2.849 | 2.673 | 2.777 | 3.084 | 3.012 |
| Proportion of 1SW fish in catch: | 0.919 | 0.919 | 0.968 | 0.941 | 0.982 | 0.968 | 0.968 | 0.968 |
| Catch of 1SW NA fish: | 16635 | 25607 | 22300 | 22392 | 17238 | 3029 | 5357 | 4712 |
| Catch of 1SW EU fish: | 13706 | 21098 | 9349 | 8000 | 4091 | 806 | 607 | 2038 |
| Natural mortality during migration: | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
|  |  |  |  |  |  |  |  |  |
| Additional fish if no fishery at Greenland: |  |  |  |  |  |  | $\mathbf{4 2 6 3}$ |  |
| 2SW fish returning to NA (numbers): | $\mathbf{1 5 0 5 2}$ | $\mathbf{2 3 1 7 1}$ | $\mathbf{2 0 1 7 7}$ | $\mathbf{2 0 2 6 2}$ | $\mathbf{1 5 5 9 8}$ | $\mathbf{2 7 4 0}$ | $\mathbf{4 8 4 7}$ | $\mathbf{4 2 6 3}$ |
| 2SW fish returning to EU (numbers): | $\mathbf{1 2 4 0 2}$ | $\mathbf{1 9 0 9 1}$ | $\mathbf{8 4 5 9}$ | $\mathbf{7 2 3 9}$ | $\mathbf{3 7 0 2}$ | $\mathbf{7 2 9}$ | $\mathbf{5 4 9}$ | $\mathbf{1 8 4 4}$ |


| Average number of salmon potentially returning to home waters per ton caught in |
| :--- |
| Greenland: |
| 2SW fish returning to NA (numbers per ton, average of 1993-2000): |
| 2SW fish returning to EU (numbers per ton, average of 1993-2000): |

${ }^{1}$ ) Figures for 1993 and 1994 correspond to calculated quotas.

Table 5.6.1.1. Pre-fishery abundance estimates, thermal habitat index for February based on sea surface temperature (H2), lagged spawner index for North America excluding Gulf and US spawners (SLNQ), results of a jacknife cross-validation of the multiplicative forecast model, and simulated forecasts.

| Year | Pre-fishery abundance |  |  | Thermal <br> Habitat <br> February (H2) | Lagged spawners (SLNQ) |  |  | Jacknife Cross-validation |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Low | High | Mid-point |  | Low | High | Mid-point | Prediction | Residuals |
| 1971 | 578,974 | 726,622 | 652,798 | 2,011 |  |  |  |  |  |
| 1972 | 557,790 | 732,940 | 645,365 | 1,990 |  |  |  |  |  |
| 1973 | 672,631 | 867,684 | 770,157 | 1,708 |  |  |  |  |  |
| 1974 | 623,907 | 800,542 | 712,224 | 1,862 |  |  |  |  |  |
| 1975 | 710,252 | 904,626 | 807,439 | 1,827 |  |  |  |  |  |
| 1976 | 610,799 | 826,787 | 718,793 | 1,676 |  |  |  |  |  |
| 1977 | 506,919 | 667,787 | 587,353 | 1,915 |  |  |  |  |  |
| 1978 | 288,792 | 371,342 | 330,067 | 1,951 | 35,453 | 81,767 | 58,610 | 389,220 | -59,153 |
| 1979 | 630,091 | 831,411 | 730,751 | 2,058 | 42,626 | 94,677 | 68,652 | 664,772 | 65,978 |
| 1980 | 550,336 | 734,489 | 642,412 | 1,823 | 43,173 | 97,017 | 70,095 | 590,190 | 52,222 |
| 1981 | 527,318 | 684,352 | 605,835 | 1,912 | 43,268 | 97,575 | 70,421 | 658,224 | -52,389 |
| 1982 | 439,982 | 567,499 | 503,741 | 1,703 | 43,381 | 98,372 | 70,876 | 563,713 | -59,972 |
| 1983 | 236,377 | 337,388 | 286,882 | 1,416 | 40,413 | 91,967 | 66,190 | 364,762 | -77,880 |
| 1984 | 245,424 | 347,471 | 296,448 | 1,257 | 37,647 | 84,066 | 60,856 | 233,165 | 63,283 |
| 1985 | 399,028 | 539,102 | 469,065 | 1,410 | 39,344 | 83,435 | 61,389 | 248,799 | 220,266 |
| 1986 | 435,090 | 575,673 | 505,381 | 1,688 | 40,567 | 91,757 | 66,162 | 442,148 | 63,233 |
| 1987 | 398,168 | 527,764 | 462,966 | 1,627 | 36,636 | 88,818 | 62,727 | 353,451 | 109,515 |
| 1988 | 317,609 | 423,746 | 370,678 | 1,698 | 37,131 | 83,891 | 60,511 | 339,966 | 30,712 |
| 1989 | 241,044 | 345,930 | 293,487 | 1,642 | 41,955 | 86,459 | 64,207 | 400,432 | -106,945 |
| 1990 | 218,191 | 296,332 | 257,262 | 1,503 | 40,948 | 81,667 | 61,307 | 304,340 | -47,078 |
| 1991 | 249,798 | 349,917 | 299,857 | 1,357 | 37,582 | 72,966 | 55,274 | 178,975 | 120,882 |
| 1992 | 143,925 | 216,262 | 180,094 | 1,381 | 35,596 | 71,384 | 53,490 | 179,100 | 994 |
| 1993 | 95,352 | 179,428 | 137,390 | 1,252 | 38,387 | 79,232 | 58,810 | 244,899 | -107,509 |
| 1994 | 110,985 | 219,159 | 165,072 | 1,329 | 38,395 | 75,762 | 57,079 | 215,540 | -50,467 |
| 1995 | 120,523 | 202,958 | 161,740 | 1,311 | 36,740 | 69,943 | 53,342 | 168,198 | -6,458 |
| 1996 | 104,675 | 163,182 | 133,928 | 1,470 | 33,492 | 61,600 | 47,546 | 134,001 | -72 |
| 1997 | 69,083 | 123,311 | 96,197 | 1,594 | 29,876 | 55,241 | 42,558 | 107,109 | -10,912 |
| 1998 | 58,751 | 126,207 | 92,479 | 1,849 | 25,629 | 50,461 | 38,045 | 91,858 | 621 |
| 1999 | 57,800 | 130,436 | 94,118 | 1,741 | 25,658 | 52,637 | 39,147 | 91,254 | 2,864 |
| 2000 |  |  |  | 1,634 | 32,960 | 68,185 | 50,572 | 225,708 ${ }^{1}$ |  |
| 2001 |  |  |  | 1,685 | 37,414 | 81,709 | 59,561 | 295,678 ${ }^{1}$ |  |

[^13]Table 5.6.2.1 Results of analysis of prefishery abundance (NN1) on February thermal habitat (H2) and North American spawners (SLNQ) from the multiplicative model, 1978-99.

## General Linear Models Procedure

Dependent Variable: LNN1

| Source | DF | Sum of <br> Squares | Mean <br> Square | F Value | Pr > |
| :--- | :---: | ---: | :--- | ---: | :--- |
| Model | 2 | 7.83182436 | 3.91591218 | 70.52 |  |

R-Square
0.881287

| Source | DF | Type I SS | Mean Square | F Value | Pr > F |
| :---: | :---: | :---: | :---: | :---: | :---: |
| H2 | 1 | 1.32987051 | 1.32987051 | 23.95 | 0.0001 |
| LN (SLNQ) | 1 | 6.50195385 | 6.50195385 | 117.10 | $<.0001$ |
| Source | DF | Type III SS | Mean Square | F Value | Pr $>\boldsymbol{F}$ |
| H2 | 1 | 0.80611073 | 0.80611073 | 14.52 | 0.0012 |
| LN(SLNQ) | 1 | 6.50195385 | 6.50195385 | 117.10 | <.0001 |

## Regression statistics

| Parameter | Estimate | Standard <br> Error | t Value | Pr $>\|t\|$ |
| :--- | ---: | ---: | ---: | ---: |
| INTERCEPT |  |  |  |  |
| H2 | -22.69738188 | 3.11673567 | -7.28 | $<.0001$ |
| LN (SLNQ) | 0.00082736 | 0.00021714 | 3.81 | 0.0012 |
|  | 3.09084662 | 0.28562819 | 10.82 | $<.0001$ |

Table 5.6.2.2 Estimate of pre-fishery abundance in 2001. forecasted by H2-SLNQ multiplicative model of probability levels between 25 and $75 \%$.
Cumulative Density

| Function \% | Forecast |
| :---: | :---: |
| 25 | 187,700 |
| 30 | 207,784 |
| 35 | 228,189 |
| 40 | 249,433 |
| 45 | 271,859 |
| 50 | 295,678 |
| 55 | 321,537 |
| 60 | 350,175 |
| 65 | 382,308 |
| 70 | 418,919 |
| 75 | 462,797 |
|  |  |

Table 5.6.3.1 Quota options (mt) for 2001 at West Greenland based on H2-SLNQ multiplicative forecasts of fishery abundance. Proportion at West Greenland refers to the fraction of harvestable surplus allocated to the West Greenland fishery. The probability level refers to the pre-fishery abundance levels derived from the probability density function.

| Prob. | Proportion at West Greenland (Fna) |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| level | 0 | 0.1 | 0.2 | 0.3 | 0.4 | 0.5 | 0.6 | 0.7 | 0.8 | 0.9 | 1 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 25 | 0 | 7 | 14 | 21 | 28 | 35 | 42 | 49 | 56 | 63 | 69 |  |  |
| 30 | 0 | 15 | 30 | 45 | 60 | 75 | 90 | 105 | 120 | 135 | 150 |  |  |
| 35 | 0 | 23 | 46 | 69 | 92 | 116 | 139 | 162 | 185 | 208 | 231 |  |  |
| 40 | 0 | 32 | 63 | 95 | 126 | 158 | 190 | 221 | 253 | 284 | 316 |  |  |
| 45 | 0 | 41 | 81 | 122 | 162 | 203 | 243 | 284 | 324 | 365 | 405 |  |  |
| 50 | 0 | 50 | 100 | 150 | 200 | 250 | 300 | 350 | 400 | 450 | 500 |  |  |
| 55 | 0 | 60 | 121 | 181 | 241 | 302 | 362 | 423 | 483 | 543 | 604 |  |  |
| 60 | 0 | 72 | 144 | 215 | 287 | 359 | 431 | 503 | 574 | 646 | 718 |  |  |
| 65 | 0 | 85 | 169 | 254 | 338 | 423 | 508 | 592 | 677 | 761 | 846 |  |  |
| 70 | 0 | 99 | 198 | 298 | 397 | 496 | 595 | 695 | 794 | 893 | 992 |  |  |
| 75 | 0 | 117 | 233 | 350 | 467 | 584 | 700 | 817 | 934 | 1,051 | 1,167 |  |  |


| Sp. res $=$ | 170,286 |
| :--- | ---: |
| Prop NA $=$ | 0.779 |
| WT1SWNA $=$ | 2.954 |
| WT1SWE $=$ | 2.990 |
| ACF $=$ | 1.049 |

Table 5.7.2.1 Estimate of pre-fishery abundance in 2001. forecasted by H2-SLNQ additive model of probability levels between 25 and 75\%.

| Cumulative Density |  |
| :---: | :---: |
|  |  |
| Function \% | Forecast |
| 25 | 312,426 |
| 30 | 324,966 |
| 35 | 336,579 |
| 40 | 347,550 |
| 45 | 358,165 |
| 50 | 368,685 |
| 55 | 379,156 |
| 60 | 389,811 |
| 65 | 400,880 |
| 70 | 412,467 |
| 75 | 424,968 |
|  |  |

Table 5.7.2.2 Quota options (mt) for 2001 at West Greenland based on H2-SLNQ additive model forecasts of fishery abundance. Proportion at West Greenland refers to the fraction of harvestable surplus allocated to the West Greenland fishery. The probability level refers to the pre-fishery abundance levels derived from the probability density function.

| Prob. | Proportion at West Greenland (Fna) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| level | 0 | 0.1 | 0.2 | 0.3 | 0.4 | 0.5 | 0.6 | 0.7 | 0.8 | 0.9 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2 | 0 | 57 | 113 | 170 | 227 | 284 | 340 | 397 | 454 | 511 |  |  |
| 30 | 0 | 62 | 123 | 185 | 247 | 309 | 370 | 432 | 494 | 556 |  |  |
| 35 | 0 | 66 | 133 | 199 | 265 | 332 | 398 | 465 | 531 | 597 |  |  |
| 40 | 0 | 71 | 141 | 212 | 283 | 354 | 424 | 495 | 566 | 637 |  |  |
| 45 | 0 | 75 | 150 | 225 | 300 | 375 | 450 | 525 | 600 | 675 |  |  |
| 50 | 0 | 79 | 158 | 238 | 317 | 396 | 475 | 554 | 633 | 713 |  |  |
| 55 | 0 | 83 | 167 | 250 | 333 | 417 | 500 | 583 | 667 | 750 |  |  |
| 60 | 0 | 88 | 175 | 263 | 350 | 438 | 526 | 613 | 701 | 788 |  |  |
| 65 | 0 | 92 | 184 | 276 | 368 | 460 | 552 | 644 | 736 | 828 |  |  |
| 70 | 0 | 97 | 193 | 290 | 387 | 483 | 580 | 677 | 773 | 870 |  |  |
| 75 | 0 | 102 | 203 | 305 | 407 | 508 | 610 | 711 | 813 | 915 |  |  |


| Sp. res $=$ | 170,286 |
| :--- | ---: |
| Prop NA $=$ | 0.779 |
| WT1SWNA $=$ | 2.954 |
| WT1SWE = | 2.990 |
| ACF $=$ | 1.049 |

Figure 5.1.2.1 Numbers of North American and European Atlantic salmon caught at West Greenland 1982-1992 and 1995-2000.


Figure 5.1.3.1. Mean weight ( kg , fresh round weight) of 1 SW Atlantic salmon of North American and European origin sampled at West Greenland from 1969 to 2000


Figure 5.3.1. The proportion (with confidence intervals) of North American salmon for (A) NAFO Divisions and (B) for years 1987-1999.


Figure 5.3.2. The proportions of North American salmon in samples at West Greenland and in the total pre-fishery abundance of North American and southern European 1SW non-maturing salmon, 1971-1999.


Figure 5.4.1 Extant exploitation of the non-maturing component of North American salmon as 1SW North America and Greenland from the run reconstruction statistics.


Fig. 5.6.1.1. Thermal habitat index for Frebruary (H2) and lagged spawners (SLNQ).


Figure 5.6.2.1. Observed estimates, jackknifed historical predictions, and deterministic forecasts (upper Panel A) of prefishery abundance from the multiplicative model. The residual pattern from the jackknifed predictions is shown in the lower panel (Panel B).


Figure 5.6.3.1. Number of fish yielding a 200 mt quota, relative to changes in the median date of the fishery and associated changes in mean weight of fish harvested. Equations presented by Jensen (1990) were used to calculate daily changes in mean weight.


Figure 5.6.4.1. Exact (upper) and cumulative (lower) posterior predicted probability distributions of the PFA in year 2001 based on the multiplicative model of survival with errors in the PFA and SNLQ variables. The distributions were generated from 50,000 Monte Carlo simulations.


Figure 5.6.4.2. Risk analysis (probability of meeting the conservation requirement simultaneously in the six stock areas in North America) of catch options on the prefishery 1SW non-maturing salmon component in 2001.


Figure 5.7.1.1. Observed estimates, jacknifed historical predictions (upper panel) of pre-fishery abundance. The residual pattern for the additive and multiplicative forecast models is shown in the lower panel.


Figure 5.8.1. Association between the juvenile index and the estimated PFA for North America, 1978 to 1999.


Figure 5.8.2. PFA predictions from the juvenile index model for 2000 (upper) and 2001 (lower).



The Working Group recommends that it should meet in 2002 to address questions posed by ACFM, including those posed by NASCO. No invitation to host the meeting was proposed to the Working Group. Therefore, the Working Group should convene from the $3^{\text {rd }}$ to the $13^{\text {th }}$ of April 2002 in Copenhagen, Denmark. It is strongly recommended by the Working Group that this period is adhered in order to provide sufficient time to adequately review and complete the report

The Working Group recognises the problems with long meetings is considering solutions to increase the efficiency and reduce the length of the meeting. A large amount of time is spent updating tables and figures from the previous year's report. It suggested that ICES establish a WGNAS web site giving the opportunity for all subgroups (NEAC, NAC, West Greenland) to update as much as possible ( tables, figures and the texts associated to them) prior to the working group meeting. The Chairman would provide directions for the updating process each year. It is also suggested that the overall size of the report be shortened by removing some long series of historical data, appendices, etc, which could be more efficiently held and made available on the web. This process should be carried out concurrently with the present arrangements recommended for the Working group meeting above. Subsequently, the overall length of the meeting can be shortened based on the time being saved by adopting the web-site procedure.

### 6.2 Data deficiencies and research needs.

## Recommendations from Section 2- Atlantic salmon in the North Atlantic Area:

1. Investigate the by-catch issue to assess the possible impact on post-smolt survival in pelagic fisheries (2.4.4.1).
2. Survey and investigat infestation rates of salmon lice and their influence on subsequent returns rates. (2.4.4.2).
3. Evaluation of the maturity schedule method particularly as it relates to the sensitivities of the survival estimates to the sex ratio values of the smolts and the assumption of equal survival of male and female salmon (2.4.6).

Recommendations from Section 3- Fisheries and Stocks from the North East Atlantic Commission Area:

1. More research into the biology of salmon in the marine phase is required. This includes the need to monitor trends in marine mortality for a wider range of stocks than at present, and identify causes for mortality. It should also include the examination of relationships between postsmolt growth and marine mortality. The use of data storage tags will significantly improve the information on the marine life history of salmon. $(2.4,3.9)$
2. Research on postsmolts in the early marine phase should be continued and expanded. This should include studies of interactions with parasites and assessments of the impact of sea lice on postsmolts. (2.4,3.9)
3. A Study Group is required to quantitatively assess the level of bycatches of postsmolts in pelagic fisheries. It is recommended that such a group should comprise both those with information relating to postsmolt distribution and those who can provide information on the activity and distribution of pelagic fisheries. $(2.4,3.9)$
4. A coordinated programme of tagging and release of farmed salmon should be undertaken to improve knowledge on the marine survival and migratory behaviour of these fish (2.4)
5. If the commercial fishery at Faroes recommences, it is recommended that biological samples from the salmon caught should be collected. Historical sample from this fishery information should continue to be analysed.(3.1)

## Recommendations from Section 4- Fisheries and Stocks from the North American Commission Area:

1. Estimates of total returns to Labrador no longer exist. There is a critical need to develop alternate methods to derive estimates of salmon returns and develop habitat-based spawner requirements in Labrador, and to monitor salmon returns in the Ungava regions of Québec. (4.2.2; 4.2.4)
2. There is a need to investigate changes in the biological characteristics (mean weight, sex ratio, sea-age composition) of returns to rivers, spawning stocks of Canadian and US rivers, and the harvest in aboriginal fisheries in Labrador.

These data and new information on measures of habitat and stock recruitment are necessary to re-evaluate existing estimates of spawner requirements in Canada and USA and for use in the run reconstruction model. (4.2.2; 4.2.3; 4.4)
3. There is a requirement for additional smolt-to-adult survival rates for wild salmon. As well, sea survival rates of wild salmon from rivers stocked with hatchery smolts should be examined to determine if hatchery return rates can be used as an index of sea survival of wild salmon elsewhere. (4.2.5)
4. Further basic research is needed on the spatial and temporal distribution of salmon and their predators at sea to assist in explaining variability in survival rates. (4.2.3; 4.2.5)
5. Return estimates for the few rivers (Annapolis, Cornwallis and Gaspareau) in SFA 22 that do contribute to distant fisheries should be developed and, when these are available, the SFA 22 spawning requirements for these rivers (476 fish) be included in the total.(4.4)
6. A consistent approach to estimating returns is needed, to incorporate broodstock, if offspring from such broodstock are stocked back into the management area from which their parents originated. (4.1.3)
7. Accounting of escaped-farmed salmon from North America indicates a high but undocumented mortality. Scale analysis of salmon captured at West Greenland indicated an infrequent appearance of escaped-farmed salmon. In order to substantiate this conclusion farmed-salmon need to be included in background genetic analysis and the data re-examined for the presence of escaped-farmed salmon of North American origin.

## Recommendations from Section 5- Atlantic Salmon in the West Greenland Commission Area:

7. Continued efforts should be made to improve the estimates of the annual catches of salmon taken for local consumption in Greenland. (5.1)
8. The mean weights, sea and freshwater ages and continent of origin are essential parameters to provide catch advice for the West Greenland fishery. As these parameters are known to vary over time, the Working Group recommends that the sampling programme be continued and closely coordinated with fishery harvest plan to be executed in West Greenland. (5.6)
9. The Working Group recommends that other indices of change, i.e. changes in age composition, size at age and sea survival, should also be included in the evaluation of the effects on European and North American stocks of the West Greenlandic management measures since 1993. (5.4)
10. The catch options for the West Greenland fishery are based almost entirely upon data taken from North American stocks (with the current exclusion of Labrador, see Section 4.6). In view of the evidence of a long-term decline in the European stock components contributing to this fishery (southern European non-maturing 1SW recruits) the Working Group emphasised the need for information from these stocks to be incorporated into the assessments as soon as possible. $(5.4,3.8)$
11. The Working Group recommends that an evaluation be conducted on the present reliability of the PFA estimate. An initial approach is to determine what fraction of the PFA estimate is directly based on catches and assessed returns (hard data), and what fraction results from less certain information such as scaling factors for potential productive habitat. (5.7)
12. Alternative models should be explored (for example different predictive variables, model formulations, univariate time series, non-parametric change-of-state analyses) to provide some index of plausibility of the quantitative forecasts. (5.6)

Further basic research is needed on the spatial/temporal distribution and migration patterns of salmon in relation to sea surface temperature and their predators at sea to assist in explaining variability in survival rates. (5.6)

## REPORT OF THE

# WORKING GROUP ON NORTH ATLANTIC SALMON <br> Part three - Appendices 

Aberdeen, Scotland

2-11 April 2001

This report is not to be quoted without prior consultation with the General Secretary. The document is a report of an expert group under the auspices of the International Council for the Exploration of the Sea and does not necessarily represent the views of the Council.

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## APPENDIX 1

WORKING DOCUMENTS SUBMITTED TO THE WORKING GROUP ON NORTH ATLANTIC SALMON, 2001
WORKING TITLES AND RELEVANT SECTIONS OF THE REPORT
Doc. No. 1 Erkinaro, J., Länsman, M., Kuusela, J., Julkunen, M. and Niemelä, E. National report for Finland: salmon fishing season in 2000

Doc. No. 2 MacLean, J.C. and G.W. Smith. National report for UK (Scotland) for the year 2000

Doc. No. 3 MacLean, J.C., Smith, G.W. and Whyte, BDM. A description of marine growth checks observed on the scales of salmon returning to Scottish homewaters between 1997 and 1999; evidence for an association between individuals in the ocean which reflects sub-catchment population structuring.

Doc. No. 4 Rowan, J., Sprankle, K., McKeon, J., Marancik, J., Rideout, S., Trial, J., Perkins and Brown, R. National Report for the United States, 2000

Doc. No. 5 Brown, R.W., Mackey, G., and Trial, J. Efforts to minimize interactions between wild and farmed salmon

Doc. No. 6 Caron , F. and Fontaine, P.M. Status of Atlantic Salmon stocks in Quebec.

Doc. No. 7 Anon. Atlantic Salmon Maritime Provinces - Overview for 2000

Doc. No. $8 \quad$ Whoriskey, F.G., Jnr. Infectious Salmon Anemia (ISA): Update on the situation in North America.

Doc. No. 9 Whoriskey, F.G., Jnr. Causes of the escape of farmed Atlantic salmon from sea cages in British Columbia and North America.

Doc. No. 10 Swansberg, E., El-Jabi, N., Chaput, G. and Caissie, D. Impact of climate change on river water temperatures and fish growth.

Doc. No. 11 Chaput, G., Caron, F., Marshall, L. and Amiro, P. Estimation of marine M for Atlantic Salmon

Doc. No. 12 Prusov, S.V., Prischepa, B.F., Krylova, S.S., Antonova, V.P and Bugaev, V.F. Atlantic Salmon fisheries and status of stocks in Russia - National report for 2000.

Doc. No. 13 Ó Maoiléidigh, N., A. Cullen, T. McDermott, N. Bond, D. McLaughlin, and G. Rogan. National Report for Ireland - The 2000 salmon season

Doc. No. 14 Ó Maoiléidigh, N., A. Cullen, T. McDermott, N. Bond, and D. McLaughlin. Review of Irish salmon aquaculture escapee data.

Doc. No. 15\&16 Holme, M., Holst, J.C, Hansen, L.P.H. and Nilsen, F. Salmon surveys in the NE Atlantic in 2000 -Post-smolt distribution and status of salmon lice investigations.

Doc. No. 17 Anon. Annual Assessment of Salmon stocks and Fisheries in England and Wales, 2000. Report prepared by CEFAS and the EA, UK.

Doc. No. 18 Potter, E.C.E. Annual updating and running instructions for the NEAC Area Pre-fishery Abundance and Conservation Limit model.

Doc. No. 19 Potter, E.C.E. Sensitivity analysis for NEAC area pre-fishery abundance assessment.
Doc. No. 20 Potter, E.C.E and Nicholson, M. A simple model to estimate biological reference points from noisy stock-recruitment data.

Hansen, L.P., Fiske, P., Holme, M., Jensen, A.J., Johnsen, B.O., Arnekleiv, J.V. and Hvidsten, N.A. Atlantic salmon; national report for Norway 2000.

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Short, P.B, Reddin, D.G., Johnson, R.W., King, T., Brown, R. and Kanneworff. Identification and characteristics of North American and European Atlantic salmon (Salmo salar L.) caught at West Greenland in 2000.

Doc. No. 24 Reddin, D.G. Return and spawner estimates Atlantic Salmon for Insular Newfoundland .
Doc. No. 25 Dempson, J.B., O’Connell, M.F., Reddin, D.G., Bourgeois, C., Mullins, C.C. and Porter , T.R. Newfoundland \& Labrador - Atlantic Salmon stock status for 2000.

Doc. No. 26 Meerburg, D. J. Catch, catch-and-released and unreported catch estimates for Atlantic Salmon in Canada, 2000.

Doc. No. 27 Kanneworff, P. The salmon fishery in Greenland 2000

Doc. No. 28 Gudbergsson, G., Antonsson, Th. Gudjonsson. National Report for Iceland - The 2000 salmon season.

Doc. No. 29 Crozier, W.W. , Kennedy, G.J.A. and Boylan, P. Summary of salmon fisheries and status of stocks in Northern Ireland for 2000.

Doc. No. 30 Caron, F. Atlantic salmon survival rate in freshwater and at sea on two index rivers, de la Trinitié and Saint-Jean, Québec.

Doc. No. 31 King, T.L., Brown, R.W. and Reddin, D.G. Multilocus Microsatellite DNA Genotypes as a Tool For Determining the Origins of Atlantic Salmon Collected in the West Greenland Subsistence Fishery

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## APPENDIX 3

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## APPENDIX 4

Example of SAS program to calculate Atlantic salmon pre-fishery abundance with an estimate of precision based on empirically derived distributions of observed patterns of pre-fishery abundance.

```
FILENAME CATCH DDE 'EXCEL Years78-01 ! R4C1:R27C14';
OPTIONS NOCENTER LINESIZE = 80;
*... DATA FOR CATCH ADVICE FOR 2001 FROM RISKVAR01.XLS ;
*<><><><>< UPDATE COLUMNS BY ONE IN FILENAME STATEMENT <><><>;
DATA CATCH;
    INFILE CATCH;
    INPUT YEAR NG1 NC1_L NC1_H NC2_L NC2_H NR2_L NR2_H NN1_L NN1_H NN1_M H2 GUS_L
GUS_H ;
GUS_M= (GUS_L+GUS_H)/2;
LN_NN1_M=LOG(NN1_M);
LN-GUS=LOG (GUS M);
PROC PRINT;
PROC REG;
MODEL LN_NN1_M = H2 LN_GUS/P R;
- <<< \overline{In 20}01, we cha}nged to risk model with varying spawner and PFA inputs >>>; 
\bullet * <<< also switched to multiplicative model for logged PFA and spawners >>>;
DATA D2; SET CATCH;
    SEED = 0;
DO SIM = 1 TO 1000;
    RAN_C1 = NC1_L + ((NC1_H - NC1_L) * RANUNI (SEED));
    RAN_C2 = NC2_L + ((NC2_H - NC2_L) * RANUNI (SEED));
    RAN_R2 = NR2_L + ((NR2_H - NR2_L) * RANUNI (SEED));
    RAN PFA = LOG((((RAN R2/.99005) + RAN C2)/.90483) + RAN C1 + NG1);
    RAN_SP = GUS_L + ((GUS_H - GUS_L) * RANUNI (SEED));
OUTPUT;
END;
PROC SORT; BY SIM;
PROC REG NOPRINT;
    BY SIM;
    ID YEAR;
    MODEL RAN_PFA = H2 LN_GUS/ P R;
    output out=predic p=pran_pfa stdi=stdi_pfa;
*<><><><>< REMEMBER TO CHANGE THE YEAR BELOW <><><><><><>;
data univ;
    set predic;
    if year=2001;
    do i=1 to 1000;
        new_pfa=pran_pfa+((stdi_pfa)*rannor(0));
        output;
    end;
run;
PROC UNIVARIATE DATA = UNIV;
            VAR NEW_PFA;
            OUTPUT OUT=D4 PCTLNAME=
        MEAN=M STD=S
        PCTLPRE=PFA
            PCTLPTS=5 10 15 20 25 30 35 40 45 50 55 60 65 70 75 80 85 90 95;
proc print;
run;
```

Appendix 5(i). Estimated numbers of 1SW salmon recruits, returns and spawners for Labrador.

| Commercial catches of small salmon |  |  | Grilse Recruits |  |  | Grilse to rivers |  | Labrador grilse spawners Angling catch subtracted SFA $1,2 \& 14 B$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | SFA 1 | SFA 2 | SFA 14B | SFA 1,2\&14 | Nfld | SFA 1,2 |  |  |  |
|  |  |  |  | Min | Max | Min | Max | Min | Max |
| *1969 | 10774 | 21627 | 6321 | 48912 | 122280 | 18587 | 65053 | 15476 | 61942 |
| *1970 | 14666 | 29441 | 8605 | 66584 | 166459 | 25302 | 88556 | 21289 | 84543 |
| *1971 | 19109 | 38359 | 11212 | 86754 | 216884 | 32966 | 115382 | 29032 | 111448 |
| *1972 | 14303 | 28711 | 8392 | 64934 | 162335 | 24675 | 86362 | 21728 | 83415 |
| *1973 | 3130 | 6282 | 1836 | 14208 | 35520 | 5399 | 18897 | 0 | 11405 |
| 1974 | 9848 | 37145 | 9328 | 71142 | 177856 | 27034 | 94619 | 24533 | 92118 |
| 1975 | 34937 | 57560 | 19294 | 141210 | 353024 | 53660 | 187809 | 49688 | 183837 |
| 1976 | 17589 | 47468 | 13152 | 98790 | 246976 | 37540 | 131391 | 31814 | 125665 |
| 1977 | 17796 | 40539 | 11267 | 87918 | 219796 | 33409 | 116931 | 28815 | 112337 |
| 1978 | 17095 | 12535 | 4026 | 42513 | 106282 | 16155 | 56542 | 13464 | 53851 |
| 1979 | 9712 | 28808 | 7194 | 57744 | 144360 | 21943 | 76800 | 17825 | 72682 |
| 1980 | 22501 | 72485 | 8493 | 130710 | 326776 | 49670 | 173845 | 45870 | 170045 |
| 1981 | 21596 | 86426 | 6658 | 144859 | 362147 | 55046 | 192662 | 49855 | 187471 |
| 1982 | 18478 | 53592 | 7379 | 100357 | 250892 | 38136 | 133474 | 34032 | 129370 |
| 1983 | 15964 | 30185 | 3292 | 62452 | 156129 | 23732 | 83061 | 19360 | 78689 |
| 1984 | 11474 | 11695 | 2421 | 32324 | 80811 | 12283 | 42991 | 9348 | 40056 |
| 1985 | 15400 | 24499 | 7460 | 59822 | 149555 | 22732 | 79563 | 19631 | 76462 |
| 1986 | 17779 | 45321 | 8296 | 90184 | 225461 | 34270 | 119945 | 30806 | 116481 |
| 1987 | 13714 | 64351 | 11389 | 112995 | 282486 | 42938 | 150283 | 37572 | 144917 |
| 1988 | 19641 | 56381 | 7087 | 104980 | 262449 | 39892 | 139623 | 34369 | 134100 |
| 1989 | 13233 | 34200 | 9053 | 71351 | 178377 | 27113 | 94896 | 22429 | 90212 |
| 1990 | 8736 | 20699 | 3592 | 41718 | 104296 | 15853 | 55485 | 12544 | 52176 |
| 1991 | 1410 | 20055 | 5303 | 33812 | 84531 | 12849 | 44970 | 10526 | 42647 |
| 1992 | 9588 | 13336 | 1325 | 29632 | 79554 | 17993 | 62094 | 15229 | 59331 |
| 1993 | 3893 | 12037 | 1144 | 33382 | 93231 | 25186 | 80938 | 22499 | 78251 |
| 1994 | 3303 | 4535 | 802 | 22306 | 63109 | 18159 | 56888 | 15228 | 53958 |
| 1995 | 3202 | 4561 | 217 | 28852 | 82199 | 25022 | 76453 | 22144 | 73575 |
| 1996 | 1676 | 5308 | 865 | 55634 | 159204 | 51867 | 153553 | 48362 | 150048 |
| 1997 | 1728 | 8025 |  | 72138 | 162610 | 66812 | 155963 | 64049 | 153200 |

Estimates are based on:
EST SMALL RETURNS - (COMM CATCH*PROP LAB ORIGIN)/EXP RATE, PROP SFAs1,2\&14B=.6-.8, SFA 1:0.36-0.42\&SFA 2:0.75-0.85(97) EXP RATE-SFAs $1,2 \& 14 \mathrm{~B}=.3-.5(69-91), .22-.39(92), .13-.25(93)$,

$$
-.10-.19(94), .07-.13(95), .04-.07(96), \text { SFA 1:0.07-0.14\&SFA 2:0.04-0.07 (97) }
$$

EST GRILSE RETURNS CORRECTED FOR NON-MATURING 1SW - (SMALL RET*PROP GRILSE), PROP GRILSE SFAs1,2\&14B=0.8-0.9 EST RET TO FRESHWATER - (EST GRILSE RET-GRILSE CATCHES)
EST GRILSE SPAWNERS = EST GRILSE RETURNS TO FRESHWATER - GRILSE ANGLING CATCHES

* Catches for 1969-73 are Labrador totals distributed into SFAs as the proportion of landings by SFA in 1974-78

Furthermore small catches in 1973 were adjusted by ratio of large:small in $1972 \& 74$ (SFA 1-1.4591, SFA 2-2.2225, SFA 14B-1.5506).

Appendix 5(ii). Estimated numbers of 2SW salmon recruits, returns and spawners for Labrador salmon stocks including west Greenland.

| Commercial catches of large salmon |  |  |  | Labrador 2SW Recruits,NF \& Greenland Labrador salmon <br> SFAs 1,2 \& 14B Labrador at Total+NF+WG |  |  |  |  | Labrador 2SW to rivers SFAs $1,2 \& 14 B$ |  | Labrador 2SW spawners SFAs 1,2 \& 14B <br> Angling catch subtracted |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Min | Max |  | Min | Max | Min | Max | Min | Max |
| *1969 | 18929 | 48822 | 10300 | 32483 | 69198 | 34280 | 80636 | 133032 | 3248 | 20760 | 2890 | 20287 |
| *1970 | 17633 | 45479 | 9595 | 30258 | 68490 | 56379 | 99561 | 154121 | 3026 | 20547 | 2676 | 20085 |
| *1971 | 25127 | 64806 | 13673 | 43117 | 97596 | 24299 | 85831 | 163577 | 4312 | 29279 | 4012 | 28882 |
| *1972 | 21599 | 55708 | 11753 | 37064 | 83895 | 59203 | 112096 | 178927 | 3706 | 25168 | 3435 | 24812 |
| *1973 | 30204 | 77902 | 16436 | 51830 | 117319 | 22348 | 96314 | 189771 | 5183 | 35196 | 4565 | 34376 |
| 1974 | 13866 | 93036 | 15863 | 50030 | 113827 | 38035 | 109433 | 200476 | 5003 | 34148 | 4490 | 33475 |
| 1975 | 28601 | 71168 | 14752 | 47715 | 107974 | 40919 | 109012 | 195006 | 4772 | 32392 | 4564 | 32119 |
| 1976 | 38555 | 77796 | 15189 | 55186 | 124671 | 67730 | 146485 | 245646 | 5519 | 37401 | 4984 | 36701 |
| 1977 | 28158 | 70158 | 18664 | 48669 | 110171 | 28482 | 97937 | 185706 | 4867 | 33051 | 4042 | 31969 |
| 1978 | 30824 | 48934 | 11715 | 38644 | 87155 | 32668 | 87816 | 157045 | 3864 | 26147 | 3361 | 25490 |
| 1979 | 21291 | 27073 | 3874 | 22315 | 50194 | 18636 | 50481 | 90267 | 2231 | 15058 | 1823 | 14528 |
| 1980 | 28750 | 87067 | 9138 | 51899 | 117530 | 21426 | 95490 | 189152 | 5190 | 35259 | 4633 | 34525 |
| 1981 | 36147 | 68581 | 7606 | 47343 | 106836 | 32768 | 100331 | 185233 | 4734 | 32051 | 4403 | 31615 |
| 1982 | 24192 | 53085 | 5966 | 34910 | 78873 | 43678 | 93497 | 156236 | 3491 | 23662 | 3081 | 23127 |
| 1983 | 19403 | 33320 | 7489 | 25378 | 57268 | 30804 | 67021 | 112531 | 2538 | 17181 | 2267 | 16824 |
| 1984 | 11726 | 25258 | 6218 | 18063 | 40839 | 4026 | 29802 | 62306 | 1806 | 12252 | 1478 | 11822 |
| 1985 | 13252 | 16789 | 3954 | 14481 | 32596 | 3977 | 24644 | 50494 | 1448 | 9779 | 1258 | 9530 |
| 1986 | 19152 | 34071 | 5342 | 24703 | 55734 | 17738 | 52991 | 97275 | 2470 | 16720 | 2177 | 16334 |
| 1987 | 18257 | 49799 | 11114 | 32885 | 74471 | 29695 | 76625 | 135970 | 3289 | 22341 | 2895 | 21821 |
| 1988 | 12621 | 32386 | 4591 | 20681 | 46789 | 27842 | 57355 | 94614 | 2068 | 14037 | 1625 | 13452 |
| 1989 | 16261 | 26836 | 4646 | 20181 | 45509 | 26728 | 55528 | 91673 | 2018 | 13653 | 1727 | 13270 |
| 1990 | 7313 | 17316 | 2858 | 11482 | 25967 | 9771 | 26158 | 46828 | 1148 | 7790 | 923 | 7493 |
| 1991 | 1369 | 7679 | 4417 | 5477 | 12467 | 7779 | 15596 | 25571 | 548 | 3740 | 491 | 3665 |
| 1992 | 9981 | 19608 | 2752 | 14756 | 37045 | 13713 | 28469 | 50758 | 2515 | 15548 | 2012 | 14889 |
| 1993 | 3825 | 9651 | 3620 | 10242 | 29482 | 6592 | 16834 | 36074 | 3858 | 18234 | 3624 | 17922 |
| 1994 | 3464 | 11056 | 857 | 11396 | 34514 | 0 | 11396 | 34514 | 5653 | 24396 | 5339 | 23981 |
| 1995 | 2150 | 8714 | 312 | 16520 | 51530 | 0 | 16520 | 51530 | 12368 | 44205 | 12006 | 43726 |
| 1996 | 1375 | 5479 | 418 | 11814 | 37523 | 4312 | 16126 | 41835 | 9113 | 32759 | 8838 | 32395 |
| 1997 | 1393 | 5550 |  | 13167 | 28647 | 3806 | 16973 | 32453 | 9384 | 23833 | 9221 | 23646 |

Estimates are based on:
EST LARGE RETURNS - (COMM CATCH*PROP LAB ORIGIN)/EXP RATE, PROP SFAs1,2\&14B=.6-.8,SFA 1: 0.64-0.72 \& SFA 2 0.88-0.95 (97);
EXP RATE-SFAs1,2\&14B=.7-.9(69-91),.58-.83(92),.38-.62(93),.29-.50(94), .15-.26(95), .13-.23(96),
SFA 1: 0.22-0.40, SFA 2: 0.16-0.28 (97)

EST 2SW RETURNS - (EST LARGE RETURNS*PROP 2SW), PROP 2SW SFA $1=.7-.9$,SFAs $2 \& 14 \mathrm{~B}=.6-.8$
WG - are North American 1SW salmon of river age 4 and older of which $70 \%$ are Labrador origin
EST RET TO FRESHWATER - (EST 2SW RET-2SW CATCHES)
EST 2SW SPAWNERS = EST 2SW RETURNS TO FRESHWATER - 2SW ANGLING CATCHES
*Catches for 1969-73 are Labrador totals distributed into SFAs as the proportion of landings by SFA in 1974-78.

Appendix 5(iii). Atlantic salmon returns to freshwater, total recruits prior to the commercial fishery and spawners summed for Salmon Fishing Area 3-14A, insular Newfoundland, 1969-2000.

|  | Small catch Small retns to rivel |  |  | Small recruits |  | Small spawner Large retns to rive |  |  |  | Large recruit Large catch |  |  | Large spawners |  | 2SW retns to rive 2SW spawners |  |  |  | 2SW recruits |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Retained | Min | Max | Min | Max | Min | Max | Min | Max | Min | Max | Retained | Min | Max | Min | Max | Min | Max | Min | Max |
| 1969 | 34944 | 109580 | 219669 | 219160 | 732230 | 74636 | 184725 | 10634 | 25631 | 35446 | 256307 | 2310 | 8324 | 23321 | 2193 | 8995 | 1383 | 7760 | 7311 | 89953 |
| 1970 | 30437 | 140194 | 281466 | 280388 | 938221 | 109757 | 251030 | 12731 | 29313 | 42435 | 293127 | 2138 | 10593 | 27175 | 3135 | 11517 | 2359 | 10340 | 10450 | 115168 |
| 1971 | 26666 | 112644 | 226129 | 225288 | 753763 | 85978 | 199463 | 9999 | 23221 | 33330 | 232208 | 1602 | 8397 | 21619 | 2388 | 8923 | 1817 | 8055 | 7959 | 89230 |
| 1972 | 24402 | 109282 | 219412 | 218564 | 731374 | 84880 | 195010 | 10368 | 23434 | 34560 | 234343 | 1380 | 8988 | 22054 | 2511 | 9003 | 2008 | 8240 | 8371 | 90031 |
| 1973 | 35482 | 144267 | 289447 | 288534 | 964822 | 108785 | 253965 | 13489 | 31645 | 44964 | 316451 | 1923 | 11566 | 29722 | 2995 | 11527 | 2283 | 10449 | 9985 | 115268 |
| 1974 | 26485 | 85216 | 170748 | 170431 | 569159 | 58731 | 144263 | 10541 | 21113 | 35137 | 211133 | 1213 | 9328 | 19900 | 1940 | 6596 | 1510 | 5942 | 6465 | 65964 |
| 1975 | 33390 | 112272 | 225165 | 224544 | 750550 | 78882 | 191775 | 11605 | 23260 | 38682 | 232596 | 1241 | 10364 | 22019 | 2305 | 7725 | 1888 | 7086 | 7684 | 77247 |
| 1976 | 34463 | 115034 | 230595 | 230068 | 768650 | 80571 | 196132 | 10863 | 21768 | 36211 | 217677 | 1051 | 9812 | 20717 | 2334 | 7698 | 2011 | 7198 | 7781 | 76982 |
| 1977 | 34352 | 110114 | 220501 | 220229 | 735004 | 75762 | 186149 | 9795 | 19624 | 32650 | 196237 | 2755 | 7040 | 16869 | 1845 | 6247 | 1114 | 5088 | 6151 | 62470 |
| 1978 | 28619 | 97375 | 195048 | 194751 | 650159 | 68756 | 166429 | 7892 | 15841 | 26307 | 158411 | 1563 | 6329 | 14278 | 1991 | 6396 | 1557 | 5712 | 6637 | 63959 |
| 1979 | 31169 | 107402 | 215160 | 214803 | 717199 | 76233 | 183991 | 5469 | 10962 | 18230 | 109619 | 561 | 4908 | 10401 | 1088 | 3644 | 98 | 3463 | 3625 | 36437 |
| 1980 | 35849 | 121038 | 242499 | 242076 | 808330 | 85189 | 206650 | 9400 | 18866 | 31335 | 188656 | 1922 | 7478 | 16944 | 2432 | 7778 | 1888 | 6925 | 8108 | 77784 |
| 1981 | 46670 | 157425 | 315347 | 314850 | 1051158 | 110755 | 268677 | 21022 | 42096 | 70074 | 420961 | 1369 | 19653 | 40727 | 3451 | 12035 | 3074 | 11442 | 11502 | 120353 |
| 1982 | 41871 | 141247 | 283002 | 282494 | 943342 | 99376 | 241131 | 9060 | 18174 | 30198 | 181736 | 1248 | 7812 | 16926 | 2914 | 9012 | 2579 | 8481 | 9714 | 90117 |
| 1983 | 32420 | 109934 | 220216 | 219868 | 734053 | 77514 | 187796 | 9717 | 19490 | 32391 | 194903 | 1382 | 8335 | 18108 | 2586 | 8225 | 2244 | 7677 | 8620 | 82253 |
| 1984 | 39331 | 130836 | 262061 | 261673 | 873537 | 91505 | 222730 | 8115 | 16268 | 27052 | 162684 | 511 | 7604 | 15757 | 2233 | 7060 | 2063 | 6800 | 7445 | 70602 |
| 1985 | 36552 | 121731 | 243727 | 243461 | 812424 | 85179 | 207175 | 3672 | 7370 | 12240 | 73702 | 0 | 3641 | 7339 | 958 | 3059 | 946 | 3042 | 3193 | 30593 |
| 1986 | 37496 | 125329 | 251033 | 250657 | 836778 | 87833 | 213537 | 7052 | 14140 | 23505 | 141400 | 0 | 6972 | 14060 | 1606 | 5245 | 1575 | 5198 | 5353 | 52445 |
| 1987 | 24482 | 128578 | 257473 | 257157 | 858244 | 104096 | 232991 | 6394 | 12817 | 21313 | 128170 | 0 | 6353 | 12776 | 1336 | 4433 | 1320 | 4409 | 4453 | 44329 |
| 1988 | 39841 | 133237 | 266895 | 266474 | 889652 | 93396 | 227054 | 6572 | 13183 | 21908 | 131832 | 0 | 6512 | 13123 | 1563 | 5068 | 1540 | 5033 | 5211 | 50681 |
| 1989 | 18462 | 60260 | 120661 | 120520 | 402203 | 41798 | 102199 | 3234 | 6482 | 10780 | 64815 | 0 | 3216 | 6463 | 697 | 2299 | 690 | 2289 | 2325 | 22992 |
| 1990 | 29967 | 99543 | 199416 | 199086 | 664721 | 69576 | 169449 | 5939 | 11909 | 19798 | 119093 | 0 | 5889 | 11859 | 1347 | 4401 | 1327 | 4372 | 4489 | 44011 |
| 1991 | 20529 | 64552 | 129308 | 129105 | 431027 | 44023 | 108779 | 4534 | 9090 | 15112 | 90896 | 0 | 4500 | 9056 | 1054 | 3429 | 1041 | 3410 | 3514 | 34291 |
| 1992 | 23118 | 118778 | 237811 | 118778 | 237811 | 95096 | 214129 | 16705 | 33463 | 16705 | 33463 | 0 | 16564 | 33322 | 3111 | 10554 | 3057 | 10474 | 3111 | 10554 |
| 1993 | 24693 | 134150 | 268550 | 134150 | 268550 | 107816 | 242217 | 8121 | 16267 | 8121 | 16267 | 0 | 7957 | 16103 | 1499 | 5094 | 1449 | 5017 | 1499 | 5094 |
| 1994 | 28959 | 95981 | 192138 | 95981 | 192138 | 66185 | 162342 | 8089 | 16216 | 8089 | 16216 | 0 | 7884 | 16010 | 1902 | 6174 | 1840 | 6077 | 1902 | 6174 |
| 1995 | 29055 | 202739 | 435153 | 202739 | 435153 | 172727 | 405141 | 16175 | 34633 | 16175 | 34633 | 0 | 15956 | 34414 | 3635 | 12592 | 3563 | 12481 | 3635 | 12592 |
| 1996 | 36715 | 257215 | 559079 | 257215 | 559079 | 218639 | 520504 | 21957 | 46706 | 21957 | 46706 | 0 | 21693 | 46442 | 4457 | 14159 | 4372 | 14028 | 4457 | 14159 |
| 1997 | 17388 | 99029 | 146050 | 99029 | 146050 | 80096 | 127116 | 15318 | 22183 | 15318 | 22183 | 0 | 14985 | 21850 | 3887 | 8355 | 3780 | 8190 | 3887 | 8355 |
| 1998 | 19672 | 146371 | 247035 | 146371 | 247035 | 124551 | 225216 | 23032 | 36266 | 23032 | 36266 | 0 | 22672 | 35906 | 5322 | 12453 | 5222 | 12295 | 5322 | 12453 |
| 1999 | 19960 | 156740 | 224959 | 156740 | 224959 | 135561 | 203780 | 21198 | 41674 | 21198 | 41674 | 0 | 20853 | 41329 | 4254 | 14262 | 4169 | 14126 | 4254 | 14262 |
| 2000 | 14709 | 116167 | 252102 | 116167 | 252102 | 99743 | 234572 | 11126 | 55412 | 11126 | 55412 | 0 | 10901 | 55187 | 2218 | 16360 | 1975 | 16021 | 2218 | 16360 |

SRR (Small returns to river) are the sum of Bay St. George small returns (Reddin \& Mullins 1996) plus Humber R small returns (Mullins \& Reddin 1996) plus small returns in SFAs 3-12 \& 14A SSR (Small recruits) $=$ SRR/(1-Exploitation rate commercial (ERC)) where ERC=0.5-0.7, 1969-91 \& ERC=0, 1992-98. SS (Small spawners) = SSR-(SC+(SR*0.1))
SC = small salmon catch retained
$\mathrm{SR}=$ small salmon catch released with assumed mortalities at $10 \%$
RL (RATIO large:small) are from counting facilities in SFAs $3-11,13 \& 14 \mathrm{~A}$, angling catches in SFA 12
LRR (Large returns to river) = SRR * RL
R (Large recruits) $=\operatorname{LRR}^{*}(1-$ Exploitation rate large (ERL)), where ERL=0.7-0.9, 1969-91; \& ERL=0, 1992-98.
LS (Large spawners) $=$ LRR-large catch retained (LC)-(0.1*large catch released)
2SW-RR (2SW returns to river ) = LRR*proportion 2SW of 0.4-0.6 for SFAs 12-14A \& 0.1-0.2 for SFAs 3-11.
2SW-S (2SW spawners) $=$ LS * proportion 2SW of 0.4-0.6 for SFAs 12-14A \& 0.1-0.2 for SFAs 3-11.
2SW-R (2SW recruits) $=$ LR * proportion $2 S W$ of $0.4-0.6$ for SFAs 12-14A\& 0.1-0.2 for SFAs 3-11.

Appendx 501 M ．Small，large，and $2 S v$ retur and spawner cstimete for SFA 16

| \％ | 5\％mall |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Rewnt |  | Wmantic |  |  |  | Whathtict |  |  | Nutane |  | 4W\％ |  |
|  | 素新， | Mav． | 者的． | Mx\％ | Wm． | H＊＊ | Wher | Hxa |  | vint |  |  | Hz |
| 639 | 3 ${ }^{3}$ |  |  |  | 24ctu | 34x | 137 |  | 312 | \％ | 9． 44 | W， | 4xa |
| 197 | 2x | 3xas |  |  | rave | 詚12 | 340 | 2xa | 近 | 76．a | 113緆 | 5980 |  |
| FTE | 3 TH | 5 FI | PWe | 3x5 | 5829 | 2145 | Ama | 17x5 | \％6x | 565 | 12x\％ | 2지제 | 713 |
| 19 | 514 | W62 | 188 | 470 | W6＊ | 4 4 Ex | W4＊4 | 116 | 4 4 | 748 |  | 34 |  |
| $1{ }^{174}$ | W | Qw |  | 34\％ | 訾 10 | \％${ }^{\text {\％}}$ |  | 4， 4 2l | 4 4 | \％ | 1048 | ＊＊＊ |  |
| \％ | 5165 | Cay | K202 |  | 720 |  | 24＊4 | \％ade |  | 64t | W\％ |  |  |
| 6程等 | 1180 | 3062 | \％${ }^{\text {W\％}}$ | 部新 | \％ |  | U\％ | ？ | W6 | Wex | ＋300\％ | Wer | 150\％ |
| 147 | 3－6 | 44x | W10 | 24浐 | 4010 | － | 539 | 13ck |  | ．10\％ | 3x ${ }^{\text {s }}$ | 47485 | 1）．${ }^{\text {The }}$ |
| 169\％ | Ex\％ | 9xis | W ${ }^{4}$ | 4＊3 | 976 | 10204 |  | 9217 | 20\％ | TVE |  | 2x． | 747 |
| 17\％ | $5{ }^{5} 5$ | 3175 | 223 | $6 \times 0$ | 3586 | W6． | 1014 | Eux | ［81 | \％ 8 E |  | FE\％ | TxT |
|  | $5 \mathrm{Ba}_{4}$ | WP6 | 3 ${ }^{\text {3 }}$ |  |  | －$x^{4}$ | 3 ${ }^{\text {Wem }}$ |  | \％ 81 | $4 \times 4$ | Hext | －145 | 殔緒 |
| 1410 | 176\％ | T10\％ | ＋8\％ | 9040 | Wato | 3 \％${ }^{\text {\％}}$ |  |  | 等 6 | W |  | 孝4\％ | Jut |
| 162 | 6\％ | \％ 6 ctas | 27 | 7x\％\％ |  | EXtM | 180 | 4288 | 208 | cos | 5\％\％ | ＊20\％ | $2{ }^{2}$ 堂 |
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| 19x | 1205 | 275\％ | 5400 | ＊63y | vemit | 27\％ | TEx | 2470 | W\％ | －10x5 | 2067 | 785 | 123\％ |
| ＋19 | 1 17 |  | ， 48 | －＋ 4 ce | Wex | \％${ }^{\text {atem }}$ | 武䜌 | U4tur | \％ 6 | 64b |  | 4，${ }^{3}$ | Wab |
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| 392 | 105． | 2xam | 446 | \％200 | 950 |  | 71\％ | \％4114 |  | $5{ }^{5146}$ | cose |  |  |
| 198 | Exim | 3511 | 373 | TEA | 1493 | 営4 | 56\％ | 馬高 | \％ 418 | ＊65 | 3197 | \％${ }^{2}$ | WEx |
| $17^{46}$ |  | 369 | 130 | 3－4\％ |  |  |  | \％19 | W60 | 4x\％ | WH6 | 3xim | TMe |
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| 130 | 748 | 1580 | 3xas | rus | TY\％ | －${ }^{\text {a }}$ | Ex2\％ | 1275 |  | EKEX | 1084 | 3\％${ }^{\text {\％}}$ | 7\％${ }^{\text {W \％}}$ |
| 198\％ | 168\％ | 16.4 | wz | \％ 6 \％ | \％ | 蛣县学 | 4ty | 尞趗 | 为酸 | 景嗉 | Cal | \＄10 | chat |
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| 1930 | 512 | 1275 | 305 | 26\％ | 342 | 780 | 3 S | 5 S | 6E\％ | $2 x^{2}$ | 470 | 162 |  |
| 200 | W5\％ | 1200 | ＊${ }^{\text {P\％}}$ | g10 | 472 | 710 | 206 | 4 Fe | \％ 5 | 3100 | 45 | 125 | 344 |










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Appendix 5(y)b. Returns and spawners of large salmon and 2SW salmon to SFA 16. Same procedure as for returns (Appendix 5 (v)b)

| Year | 25W epawners in SFA. 16 |  | Large Saimon Spawners to the Miramichi Rimer |  |  |  |  |  | Spawnere of large ealmon in SFA. 18 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Large spawners | $\begin{aligned} & 0.8 \\ & \text { Min. } \end{aligned}$ | $\begin{aligned} & 1.33 \\ & \text { Mex } \end{aligned}$ | Prop. 25W | 2SW Spawnere |  |  |  |
|  | Min. | Max |  |  |  |  | Min | Max | Min | Mex |
| 1971 | 3508 | 5832 | 4347 | 3478 | 5782 | 0.918 | 3192 | 5307 | 3022 | 6353 |
| 1972 | 14998 | 24924 | 17671 | 14137 | 23502 | 0.965 | 13643 | 22681 | 15535 | 25827 |
| 1973 | 17134 | 28486 | 20349 | 16279 | 27064 | 0.958 | 15592 | 25922 | 17889 | 29741 |
| 1974 | 27495 | 45711 | 34445 | 27556 | 45812 | 0.908 | 25021 | 41597 | 30281 | 50343 |
| 1975 | 16366 | 27209 | 21448 | 17158 | 28526 | 0.868 | 14893 | 24760 | 18855 | 31347 |
| 1976 | 10760 | 17889 | 14332 | 11466 | 19062 | 0.854 | 9792 | 16279 | 12500 | 20947 |
| 1977 | 27404 | 45560 | 32917 | 26334 | 43780 | 0.947 | 24938 | 41459 | 28938 | 48109 |
| 1978 | 8197 | 13627 | 10829 | 8663 | 14403 | 0.861 | 7459 | 12401 | 9520 | 15827 |
| 1979 | 2751 | 4573 | 4541 | 3633 | 6040 | 0.689 | 2503 | 4161 | 3992 | 6637 |
| 1980 | 15762 | 26204 | 18873 | 15098 | 25101 | 0.95 | 14343 | 23846 | 16592 | 27584 |
| 1981 | 2702 | 4492 | 4508 | 3686 | 6129 | 0.667 | 2459 | 4088 | 4051 | 6735 |
| 1982 | 9429 | 15676 | 13258 | 10606 | 17633 | 0.809 | 8581 | 14265 | 11655 | 19377 |
| 1983 | 5986 | 9951 | 8458 | 6766 | 11249 | 0.805 | 5447 | 9056 | 7436 | 12362 |
| 1984 | 12189 | 20264 | 14687 | 11750 | 19534 | 0.944 | 11092 | 18440 | 12912 | 21465 |
| 1985 | 15390 | 25586 | 20122 | 16098 | 26762 | 0.87 | 14005 | 23283 | 17690 | 29409 |
| 1986 | 22659 | 37670 | 30216 | 24173 | 40187 | 0.853 | 20619 | 34280 | 26564 | 44162 |
| 1987 | 12635 | 21006 | 18056 | 14445 | 24014 | 0.796 | 11498 | 19116 | 15873 | 26390 |
| 1988 | 15050 | 25021 | 20980 | 16784 | 27903 | 0.816 | 13696 | 22769 | 18444 | 30663 |
| 1989 | 8921 | 14831 | 15540 | 12432 | 20668 | 0.653 | 8118 | 13496 | 13662 | 22712 |
| 1990 | 14940 | 24838 | 27588 | 22070 | 36692 | 0.616 | 13595 | 22602 | 24253 | 40321 |
| 1991 | 15472 | 25721 | 29089 | 23271 | 38688 | 0.605 | 14079 | 23406 | 25573 | 42515 |
| 1992 | 18984 | 27603 | 35927 | 29281 | 42573 | 0.590 | 17275 | 25118 | 32176 | 46784 |
| 1993 | 21755 | 31632 | 34702 | 28282 | 41122 | 0.7 | 19797 | 28785 | 31079 | 45189 |
| 1994 | 14207 | 37140 | 27147 | 17808 | 46553 | 0.726 | 12929 | 33797 | 19569 | 51157 |
| 1995 | 18345 | 47600 | 32093 | 19188 | 49769 | 0.87 | 16694 | 43316 | 21085 | 54713 |
| 1996 | 12510 | 23804 | 23478 | 16741 | 31855 | 0.68 | 11384 | 21661 | 18397 | 35005 |
| 1997 | 10319 | 19411 | 17596 | 13357 | 25127 | 0.703 | 9390 | 17664 | 14678 | 27612 |
| 1998 | 4077 | 6795 | 9215 | 7275 | 12125 | 0.51 | 3710 | 6184 | 7995 | 13324 |
| 1999 | 8499 | 19211 | 15714 | 11543 | 26093 | 0.67 | 7734 | 17482 | 12685 | 28574 |
| 2000 | 9640 | 21238 | 17654 | 12901 | 28421 | 0.68 | 8773 | 19326 | 14177 | 31232 |

Appendis Sive．Keturns of small sabmon and $15 w$ salmon to SFA 18

| 7 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | \％ | 1．30 | \％${ }^{\text {a }}$ | 1椪 |
|  | W䜌 | Hes | Semat | \％ | Wex | H景 |  |
| \％${ }^{2}$ | W102 | K4x |  | 58453 | ＊ 84.45 | 67綡 | $4{ }^{4}+3$ |
|  | 31791 | \＄73］${ }^{\text {P3 }}$ | 蛀等等 |  | W184 |  | M159\％ |
| 钽等 |  |  |  | 新委䢒 |  |  |  |
|  |  |  | 73114 | 9723 | W764 |  | 37878 |
| 早等 |  | y |  |  | 3030 |  | 5xay |
|  | 7159 |  | 部等新 | 7ata | 3 | 710 |  |
| 䅗部 |  | $4{ }^{4} 4$ | \％ata |  | 640 | 等发离 | 363040 |
|  | Wumb |  |  | verem |  |  | 3434 |
| 㭏教量 | 4insel |  | Exyd |  | 3］ | 36䜌 | b6uk |
|  | 34030 | 30760 | \％${ }^{\text {a }}$ 相 | 3超部 | 4 ${ }^{\text {a }}$ |  |  |
| tete | 541 | 38 | 紋䜌盄 |  |  |  | N4tet |
| W We |  |  |  | 4．4．0 |  |  | － |
| 30\％ | 发教4 | 36 | 3610 | \％ill | 3数教 | 緒経 | $3 \times 18$ |
| 5 | 4030 |  | Wav | 2ax | 318010 | 300 | 0ustu |
| 1＊＊ | $51{ }^{4}$ | （1）${ }^{\text {cie }}$ | 3 | 4064 | athen |  | 803 ${ }^{1 / 4}$ |
|  |  |  | 7界教 |  |  |  |  |
| \％${ }^{\text {P／}}$ | 成䜌 |  | 縎教 |  | Thex |  |  |
| ＋＊＊＊ |  |  |  | W |  | 戓紜變 |  |
|  |  |  | Prex ${ }^{\text {a }}$ | 3013 | mbim |  | M17x |
| \％${ }^{\text {\％}}$ | 7 71Ex |  | 莫4．4 | 等亲新 | Whax |  |  |
| \％ 3 \％ |  | 380． | 10464 | 495 | 419\％ | W74＊ | n¢756 |
|  | 䍃宯？ |  |  |  |  |  |  |
| ＋3 ${ }^{\text {che }}$ | 307］ | \％${ }^{\text {a }}$ | 54ly |  |  |  | ＊07304＊ |
|  |  | 3x ${ }^{\text {a }}$ | 424x |  |  | －4xa | 13x |
|  | ， |  |  | Hesy | $3{ }^{1} 4$ |  | 3 ${ }^{\text {a }}$ |
| 虽緒 | 如显童 |  | （4） |  |  | 364 |  |
| ＋3\％ | 1030 |  | Wersu |  | W1tem |  | 50．${ }^{2}$ |
|  |  |  |  |  | nhed |  | ntim |
|  |  |  | 300 | Yum | 310 | 20． | 3\％${ }^{\text {a }}$（1） |
| \％${ }^{\text {cke }}$ |  | $4{ }^{4} \times$ | 3xa | Sum | ＋270 | 教緒 | ＊＊200 |









Appendix $5(\mathrm{v})$ d. Spawners of small salmon and 1SW salmon to SFA 16

| Year | 15W spewners to SFA 16 |  | Spawners to the Mramichi Piver |  |  | 1SWSpammers to Miramichi |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 0.8 | 1.33 | 097 | 1.00 |
|  | Mn | Mex | Smell | Min | Max | Min | Max |
| 1971 | 18714 | 32075 | 21946 | 17557 | 29188 | 17030 | 29188 |
| 1972 | 23139 | 39659 | 27135 | 21708 | 36090 | 21057 | 36090 |
| $19 / 3$ | 26169 | 4486 | \$1038 | 24550 | 40415 | 2.8514 | 4u415 |
| 1974 | 47050 | 80656 | 55186 | 44149 | 73397 | 42824 | 73397 |
| 1975 | 41332 | 70839 | 48469 | 38775 | 64454 | 37612 | 64464 |
| 1976 | 53194 | 91171 | 62380 | 49904 | 82965 | 48407 | 82965 |
| 1977 | 11296 | 19361 | 13247 | 10598 | 17619 | 10280 | 17619 |
| 1978 | 12239 | 20977 | 14353 | 11482 | 19089 | 11138 | 19089 |
| 1979 | 26305 | 45086 | 30848 | 24678 | 41028 | 23938 | 41028 |
| 1960 | 22934 | 39307 | 26894 | 21515 | 35769 | 20870 | 35769 |
| 1981 | 34049 | 58358 | 39929 | 31943 | 53106 | 30985 | 53106 |
| 1962 | 47754 | 81846 | 56000 | 44800 | 74480 | 43456 | 74480 |
| 1983 | 12662 | 21702 | 14849 | 11879 | 19749 | 11523 | 19749 |
| 1984 | 16142 | 27665 | 18929 | 15143 | 25176 | 14689 | 25176 |
| 1985 | 35658 | 61114 | 41815 | 33452 | 55614 | 32448 | 55614 |
| 1986 | 76234 | 130659 | 89398 | 71518 | 118899 | 69373 | 118899 |
| 1997 | 53533 | 91751 | 62777 | 50222 | 93493 | 40715 | 83493 |
| 1988 | 76984 | 131945 | 90278 | 72222 | 120070 | 70056 | 120070 |
| 1989 | 41250 | 70717 | 48385 | 38708 | 64352 | 37547 | 54352 |
| 1990 | 50759 | 86997 | 59524 | 47619 | 7916 ? | 46191 | 79167 |
| 1991 | 41161 | 70547 | 48269 | 38615 | 64198 | 37457 | 64198 |
| 1998 | 112317 | 168359 | 129288 | 105370 | 153206 | 102209 | 153206 |
| 1993 | 86385 | 99509 | 76416 | 62279 | 90553 | 60811 | 90553 |
| 1994 | 27829 | 75289 | 42479 | 26108 | 68513 | 25325 | 68513 |
| 1995 | 13079 | 53561 | 34084 | 12270 | 48740 | 11500 | 48740 |
| 1996 | 19278 | 51818 | 24812 | 18086 | 47154 | 17543 | 47154 |
| 1997 | 8762 | 22609 | 12979 | 8220 | 20574 | 7973 | 20574 |
| 1996 | 19347 | 29736 | 21780 | 18150 | 27060 | 17606 | 27060 |
| 1999 | 14774 | 23281 | 16962 | 13860 | 21166 | 13444 | 21186 |
| 2000 | 21105 | 30534 | 23496 | 19800 | 27766 | 19606 | 27786 |

Same procedure for escapaments as used to calculate returns.
Assumes exploitation rates of $3 \%$ for large and $34 \%$ for smal solmon for the years 1998 to 2000 . These are everage rates for 1993 to 1997 as per assessment For 1999 notve removals $=2526$.

## Appendix 5(v). Estimated Alartic sulmon raturning recruits and spiewaurs to the Mornl Rires, SFA. 17, 1970-2000.

PEI sommercial landings are also given.

| Year | Simsll recruts |  | Small spanmers |  | Large recruhb |  | Large spawners |  | 25W recruits |  | 25W spammers |  | PEI comm. tamech (nos.) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Min | Max | Min | Max | Min | Max | Mn | Max | Min | Max | Mn | Max |  |
| 1970 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 29 |
| 1972 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 305 |
| 1973 | 5 | 9 | 3 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 206 |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 396 |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 345 |
| 1976 | 14 | 28 | 8 | 22 | 2 | 5 | 1 | 4 | 2 | 5 | 1 | 4 | 573 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 606 |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\mathrm{N} / \mathrm{A}$ |
| 1979 | 2 | 5 | 1 | 4 | 5 | 9 | 3 | 7 | 5 | 9 | 3 | 7 | 45.4 |
| 1980 | 12 | 23 | 7 | 18 | 2 | 5 | 1 | 4 | 2 | 5 | 1 | 4 | 1007 |
| 1981 | 259 | 408 | 151 | 390 | 40 | 77 | 36 | 73 | \$0 | 77 | 36 | 73 | 217 |
| 1962 | 175 | 536 | 102 | 263 | 16 | 31 | 8 | 23 | 16 | 31 | 8 | 23 | 416 |
| 1983 | 17 | 32 | 10 | 25 | 17 | 32 | 15 | 30 | 17 | 32 | 15 | 30 | 326 |
| 1984 | 17 | 32 | 10 | 25 | 13 | 25 | 13 | 26 | 13 | 26 | 13 | 26 | 46 |
| 1985 | 113 | 217 | 66 | 170 | 8 | 15 | 8 | 15 | 8 | 15 | 8 | 15 |  |
| 1906 | 566 | 1008 | 390 | 852 | 5 | 11 | 5 | 11 | 5 | 11 | 5 | 11 |  |
| 1987 | 1141 | 2194 | 665 | 1718 | 66 | 128 | 66 | 128 | 66 | 128 | 66 | 128 |  |
| 1988 | 1542 | 2563 | 890 | 2320 | 96 | 188 | 96 | 185 | 96 | 185 | 96 | 185 |  |
| 1989 | 400 | 770 | 233 | 603 | 149 | 287 | 149 | 287 | 149 | 287 | 149 | 287 |  |
| 1990 | 1842 | 3539 | 1074 | 2771 | 284 | 585 | 284 | 545 | 284 | 545 | 284 | 545 |  |
| 1991 | 1576 | 3008 | 919 | 2371 | 108 | 361 | 120 | 361 | 103 | 361 | 108 | 361 |  |
| 1992 | 1873 | 3589 | 1092 | 2818 | 96 | 163 | 95 | 183 | 95 | 183 | 96 | 183 |  |
| 1903 | 1277 | 2454 | 745 | 1502 | 22 | 43 | 22 | 43 | 22 | 43 | 22 | 43 |  |
| 1994 | 200 | 383 | 117 | 291 | 168 | 300 | 165 | 306 | 108 | 309 | 165 | 305 |  |
| 1996 | 1008 | 1914 | sas | 1441 | 55 | 154 | 81 | 151 | 55 | 154 | \$1 | 151 |  |
| 1996 | 1161 | 2576 | 738 | 2154 | 159 | 351 | 154 | 347 | 158 | 351 | 154 | 347 |  |
| 1997 | 485 | 832 | 283 | 730 | 31 | 59. | 50 | 58 | 31 | 59 | 50 | 58 |  |
| 1998 | 635 | 1221 | 370 | S56 | 79 | 151 | 76 | 149 | 79 | 151 | 76 | 149 |  |
| 1998 | 377 | 728 | 221 | 570 | 23 | 585 | 20 | 41 | 23 | 45 | 20 | 41 |  |
| 2000 | 307 | 591 | 179 | 463 | 57 | 109 | 56 | 108 | 57 | 109 | 56 | 108 |  |
| 70.89 x | 213 | 410 | 124 | 321 | 21 | 40 | 20 | 40 | 21 | 40 | 20 | 40 |  |
| 90-00x | 588 | 1506 | 575 | 1499 | 108 | 210 | 106 | 208 | 108 | 270 | 106 | 208 |  |

Nates

For 1970-1900, percent small is calculated from numbers of smal and large salmon in the retsined eatch in each year. For 1981-1997 and 1999. pervent small is calculuted from mumbers of smal and large solmun taken st the Learafis Pond trap
For 1098 and 2000 , peresut small is takan foum seicing catches at Mouseys Puol
Small reeruits are culculated as small retained salmanjexploikation rate. Angler explozation was calculated as $0.34,0.347$, and 0.204 of estimated returns in 1994, 1956, and 1996, respectivety. Fer other years the mean of these values is used. The min add max

Smal spawners $=$ number of small recruits - number of small retsined
Large rescuits = (mumber af small reeruits/(0.01'percent small)-aumber of smal reczuits
Large spamiers " namber of large recouins - number of large retained
it is asssumed thas large salmon and $2 S W$ salmon are equivilent

Appendix 5(viia). Total 2SW returns and spawners to SFA 18, 1070-2000.

|  | LARGE RETURNS |  |  |  | Commercial catches TOTAL 2SW |  |  |  |  |  |  | LARGE SPAWNERS |  |  |  | TOTAL 2SW |  |  | SFA |  | Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Margaree Large salmon |  | SFA 18 |  | 2SW RETURNS |  |  | 2SW ctch |  | RETURNS (inc. comm.) |  | Margaree |  | SFA 18 |  | SPAWNERS |  |  |  |  |  |
|  |  |  | 1.24 | 2.28 | 0.77 | 0.87 | Zone 6 | 0.77 | 0.87 |  |  | 1.24 | 2.28 | 0.77 | 0.87 |  |  |  |  |
| Year | MIN | MAX | MIN | MAX | MIN | MAX | (kg) | MIN | MAX | MIN | MAX |  |  | MIN | MAX | MIN | MAX |  |  |  |  | MIN | MAX |
| 1970 | 581 | 1,000 | 723 | 2,285 | 556 | 1,988 | 30,440 | 4,262 | 4,815 | 4,818 | 6,803 | 657 | 1,145 | 817 | 2,616 | 629 | 2,276 | Yr | 18.00 | aree |  |
| 1971 | 254 | 437 | 316 | 999 | 243 | 869 | 12,001 | 1,680 | 1,898 | 1,923 | 2,767 | 256 | 446 | 318 | 1,019 | 245 | 887 | 84 | 449 | 305 | 1.47 |
| 1972 | 284 | 488 | 353 | 1,116 | 272 | 971 | 31,840 | 4,458 | 5,037 | 4,729 | 6,007 | 272 | 474 | 338 | 1,083 | 261 | 942 | 85 | 1706 | 1215 | 1.40 |
| 1973 | 316 | 544 | 393 | 1,243 | 303 | 1,082 | 27,694 | 3,877 | 4,381 | 4,180 | 5,462 | 287 | 499 | 356 | 1,141 | 274 | 993 | 86 | 4448 | 2636 | 1.69 |
| 1974 | 289 | 498 | 360 | 1,137 | 277 | 989 | 37,437 | 5,241 | 5,922 | 5,518 | 6,911 | 318 | 554 | 396 | 1,265 | 305 | 1,101 | 87 | 3012 | 1857 | 1.62 |
| 1975 | 173 | 298 | 215 | 681 | 166 | 592 | 23,631 | 3,308 | 3,738 | 3,474 | 4,330 | 214 | 372 | 266 | 850 | 205 | 740 | 88 | 3078 | 1932 | 1.59 |
| 1976 | 222 | 381 | 276 | 871 | 213 | 757 | 18,361 | 2,571 | 2,904 | 2,783 | 3,662 | 267 | 465 | 332 | 1,062 | 256 | 924 | 89 | 3206 | 1570 | 2.04 |
| 1977 | 378 | 651 | 470 | 1,487 | 362 | 1,294 | 26,221 | 3,671 | 4,148 | 4,033 | 5,442 | 393 | 683 | 488 | 1,561 | 376 | 1,358 | 90 | 2391 | 1507 | 1.59 |
| 1978 | 427 | 735 | 531 | 1,679 | 409 | 1,461 | 30,216 | 4,230 | 4,780 | 4,639 | 6,241 | 510 | 888 | 635 | 2,028 | 489 | 1,765 | 91 | 3470 | 1757 | 1.97 |
| 1979 | 219 | 377 | 272 | 861 | 210 | 749 | 7,917 | 1,108 | 1,252 | 1,318 | 2,002 | 265 | 461 | 330 | 1,053 | 254 | 916 | 92 | 3315 | 1938 | 1.71 |
| 1980 | 378 | 651 | 470 | 1,487 | 362 | 1,294 | 24,412 | 3,418 | 3,862 | 3,780 | 5,156 | 497 | 865 | 618 | 1,976 | 476 | 1,719 | 93 | 2372 | 1102 | 2.15 |
| 1981 | 375 | 647 | 466 | 1,478 | 359 | 1,286 | 15,562 | 2,179 | 2,462 | 2,538 | 3,748 | 451 | 785 | 561 | 1,793 | 432 | 1,560 | 94 | 2043 | 1479 | 1.38 |
| 1982 | 484 | 833 | 602 | 1,903 | 463 | 1,656 | 26,664 | 3,733 | 4,218 | 4,196 | 5,874 | 555 | 965 | 690 | 2,205 | 531 | 1,919 | 95 | 1633 | 1060 | 1.54 |
| 1983 | 402 | 693 | 500 | 1,583 | 385 | 1,378 | 24,280 | 3,399 | 3,841 | 3,784 | 5,218 | 480 | 834 | 596 | 1,906 | 459 | 1,659 | 96 | 3921 | 1864 | 2.10 |
| 1984 | 327 | 583 | 407 | 1,332 | 313 | 1,159 | 15,140 | 2,120 | 2,395 | 2,433 | 3,554 | 296 | 532 | 368 | 1,216 | 283 | 1,058 | 97 | 2609 | 2098 | 1.24 |
| 1985 | 1,109 | 2,217 | 1,379 | 5,065 | 1,062 | 4,407 |  | 0 | 0 | 1,062 | 4,407 | 1,025 | 2,133 | 1,275 | 4,874 | 981 | 4,240 | 98 | 2,163 | 1,327 | 1.63 |
| 1986 | 2,738 | 5,680 | 3,405 | 12,978 | 2,622 | 11,291 |  | 0 | 0 | 2,622 | 11,291 | 2,583 | 5,525 | 3,212 | 12,624 | 2,473 | 10,983 | 99 | 1,853 | 811 | 2.28 |
| 1987 | 2,976 | 6,540 | 3,701 | 14,943 | 2,850 | 13,000 |  | 0 | 0 | 2,850 | 13,000 | 2,860 | 6,424 | 3,557 | 14,678 | 2,739 | 12,770 | 0 | 959 | 636 | 1.51 |
| 1988 | 1,286 | 2,494 | 1,599 | 5,698 | 1,231 | 4,958 |  | 0 | 0 | 1,231 | 4,958 | 1,143 | 2,351 | 1,421 | 5,372 | 1,094 | 4,673 |  |  | Min | 1.244 |
| 1989 | 1,708 | 3,693 | 2,124 | 8,438 | 1,635 | 7,341 |  | 0 | 0 | 1,635 | 7,341 | 1,583 | 3,568 | 1,969 | 8,152 | 1,516 | 7,092 |  |  | Max | 2.285 |
| 1990 | 3,481 | 7,933 | 4,329 | 18,126 | 3,333 | 15,769 |  | 0 | 0 | 3,333 | 15,769 | 3,347 | 7,799 | 4,162 | 17,819 | 3,205 | 15,503 |  |  |  |  |
| 1991 | 1,853 | 5,785 | 2,304 | 13,218 | 1,774 | 11,499 |  | 0 | 0 | 1,774 | 11,499 | 1,692 | 5,624 | 2,104 | 12,850 | 1,620 | 11,179 |  |  |  |  |
| 1992 | 4,875 | 9,375 | 6,062 | 21,420 | 4,668 | 18,636 |  | 0 | 0 | 4,668 | 18,636 | 4,722 | 9,222 | 5,872 | 21,071 | 4,522 | 18,332 |  |  |  |  |
| 1993 | 2,408 | 6,158 | 2,995 | 14,070 | 2,306 | 12,241 |  | 0 | 0 | 2,306 | 12,241 | 2,274 | 6,024 | 2,828 | 13,764 | 2,177 | 11,975 |  |  |  |  |
| 1994 | 2,350 | 4,500 | 2,922 | 10,282 | 2,250 | 8,945 |  | 0 | 0 | 2,250 | 8,945 | 2,209 | 4,359 | 2,747 | 9,960 | 2,115 | 8,665 |  |  |  |  |
| 1995 | 1,750 | 3,815 | 2,176 | 8,717 | 1,676 | 7,583 |  | 0 | 0 | 1,676 | 7,583 | 1,693 | 3,758 | 2,105 | 8,586 | 1,621 | 7,470 |  |  |  |  |
| 1996 | 2,214 | 4,050 | 2,753 | 9,254 | 2,120 | 8,051 |  | 0 | 0 | 2,120 | 8,051 | 2,001 | 3,837 | 2,488 | 8,767 | 1,916 | 7,627 |  |  |  |  |
| 1997 | 3,268 | 5,435 | 4,064 | 12,418 | 3,129 | 10,804 |  | 0 | 0 | 3,129 | 10,804 | 3,006 | 5,173 | 3,738 | 11,819 | 2,878 | 10,283 |  |  |  |  |
| 1998 | 2,283 | 3,798 | 2,839 | 8,678 | 2,186 | 7,550 |  | 0 | 0 | 2,186 | 7,550 | 2,114 | 3,629 | 2,629 | 8,292 | 2,024 | 7,214 |  |  |  |  |
| 1999 | 1,440 | 2,400 | 1,791 | 5,484 | 1,379 | 4,771 |  | 0 | 0 | 1,379 | 4,771 | 1,276 | 2,236 | 1,587 | 5,109 | 1,222 | 4,445 |  |  |  |  |
| 2000 | 991 | 1648 | 1,232 | 3,765 | 949 | 3,276 |  | 0 | 0 | - 949 | 3,276 | 776 | 1,433 | 965 | 3,274 | 743 | 2,849 |  |  |  |  |

Margaree returns, 1970-84, equal catch $/ \min (0.215)$ or $\max (0.37)$ exploitation rate.
Return of large salmon (MIN) and (MAX) to all SFA 18 equals Margaree returns * ratio Margaree catch to SFA 18 catch.
Margaree returns 1984-00 based on various Margaree CAFSAC , DFO Atl. Res and CSAS Res. Docs.
Margaree catch to SFA 18 catch; MIN _MAX 2SW based on the ratio 0.77-0.87 2SW fish among MSW fish.
Margaree escapements 1970-83 = returns minus removals; 1984-1999 from various Margaree CAFSAC, DFO Atl. Fish. and CSAS Res.
Docs e.g., Marshall et al. (MS 1997) where 2SW equal 0.77-0.87 of MSW fish; Margaree raised to SFA by respective ratios in recreational catch.

# Appendix 5(viib). Total 1SW returns and spawners to SFA 18, 1970-2000. 

| Year | RETURNS |  |  |  | SPAWNERS |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Margaree |  | SFA 18 |  | Margaree |  | SFA 18 |  |
|  | 0.37 | 0.21 | 1.214 | 2.768 |  |  | 1.214 | 2.768 |
|  | MIN | MAX | MIN | MAX | MIN | MAX | MIN | MAX |
| 1970 | 230 | 395 | 279 | 1,094 | 145 | 310 | 176 | 859 |
| 1971 | 57 | 98 | 69 | 270 | 36 | 77 | 43 | 212 |
| 1972 | 114 | 195 | 138 | 541 | 72 | 153 | 87 | 424 |
| 1973 | 449 | 772 | 545 | 2,137 | 283 | 606 | 343 | 1,678 |
| 1974 | 162 | 279 | 197 | 772 | 102 | 219 | 124 | 606 |
| 1975 | 97 | 167 | 118 | 463 | 61 | 131 | 74 | 364 |
| 1976 | 259 | 447 | 315 | 1,236 | 163 | 351 | 198 | 970 |
| 1977 | 186 | 321 | 226 | 888 | 117 | 252 | 143 | 697 |
| 1978 | 68 | 116 | 82 | 322 | 43 | 91 | 52 | 253 |
| 1979 | 1,614 | 2,777 | 1,959 | 7,685 | 1,017 | 2,180 | 1,234 | 6,033 |
| 1980 | 451 | 777 | 548 | 2,150 | 284 | 610 | 345 | 1,688 |
| 1981 | 2,430 | 4,181 | 2,950 | 11,573 | 1,531 | 3,282 | 1,859 | 9,085 |
| 1982 | 1,868 | 3,214 | 2,267 | 8,896 | 1,177 | 2,523 | 1,429 | 6,983 |
| 1983 | 184 | 316 | 223 | 875 | 116 | 248 | 141 | 687 |
| 1984 | 400 | 688 | 486 | 1,904 | 158 | 446 | 192 | 1,234 |
| 1985 | 634 | 1,167 | 770 | 3,230 | 125 | 658 | 152 | 1,821 |
| 1986 | 838 | 1,420 | 1,017 | 3,930 | 56 | 638 | 68 | 1,766 |
| 1987 | 1,143 | 1,865 | 1,388 | 5,162 | 166 | 888 | 202 | 2,458 |
| 1988 | 1,674 | 2,911 | 2,032 | 8,057 | 795 | 2,032 | 965 | 5,624 |
| 1989 | 591 | 977 | 718 | 2,704 | 30 | 416 | 36 | 1,151 |
| 1990 | 940 | 5,077 | 1,141 | 14,052 | 291 | 4,428 | 353 | 12,256 |
| 1991 | 794 | 3,891 | 964 | 10,770 | 42 | 3,139 | 51 | 8,688 |
| 1992 | 1,258 | 2,419 | 1,527 | 6,695 | 701 | 1,862 | 851 | 5,154 |
| 1993 | 1,489 | 3,851 | 1,808 | 10,659 | 906 | 3,268 | 1,100 | 9,045 |
| 1994 | 573 | 1,101 | 696 | 3,047 | 259 | 787 | 314 | 2,178 |
| 1995 | 538 | 1,083 | 653 | 2,998 | 329 | 874 | 399 | 2,419 |
| 1996 | 1,277 | 2,960 | 1,550 | 8,193 | 935 | 2,618 | 1,135 | 7,246 |
| 1997 | 316 | 1,517 | 384 | 4,199 | 68 | 1,269 | 83 | 3,512 |
| 1998 | 349 | 1,625 | 424 | 4,498 | 126 | 1,402 | 153 | 3,880 |
| 1999 | 323 | 1,610 | 392 | 4,456 | 100 | 1,387 | 121 | 3,839 |
| 2000 | 223 | 1,368 | 271 | 3,786 | 84 | 1,229 | 102 | 3,402 |


| Recreational ctch: |  |  |  |
| ---: | ---: | ---: | ---: |
| Year | SFA 18 | Marg- <br> aree | Ratio |
| 1984 | 298 | 242 | 1.23 |
| 1985 | 618 | 509 | 1.21 |
| 1986 | 1,180 | 782 | 1.51 |
| 1987 | 1,289 | 977 | 1.32 |
| 1988 | 1,349 | 879 | 1.53 |
| 1989 | 928 | 561 | 1.65 |
| 1990 | 1,206 | 649 | 1.86 |
| 1991 | 1,262 | 752 | 1.68 |
| 1992 | 1,242 | 678 | 1.83 |
| 1993 | 1,216 | 777 | 1.56 |
| 1994 | 659 | 429 | 1.54 |
| 1995 | 711 | 333 | 2.14 |
| 1996 | 2,022 | 918 | 2.20 |
| 1997 | 558 | 316 | 1.77 |
| 1998 | 829 | 349 | 2.38 |
| 1999 | 894 | 323 | 2.77 |
| 2000 | 467 | 223 | 2.09 |
|  |  |  |  |
|  |  | Min | 1.214 |
|  |  | Max | 2.768 |

Margaree returns, 1970-1983, equal catch divided by MIN ( 0.37 ) and MAX ( 0.215 ) exploitation rate.
Return of small salmon to all SFA 18 equals Margaree returns * MIN and MAX ratio of
Margaree catch to SFA 18 catch. Margaree returns, 1984-2000, based on annual assessments
in CAFSAC and DFO Atl. Fish. and CSAS Res. Docs, eg., Marshall et al. (MS 1997).
Spawners for 1970-1983 equal returns minus removals; 1984-2000 from various Margaree CAFSAC,
Atl. Res. and CSAS Res Doc. series.

Appendix 5(viii). Total 1SW returns and spawners, SFAs 19, 20, 21 and 23, 1970-2000.

| Year | RETURNS |  |  |  |  |  | TOTAL RETURNS |  | SPAWNERS |  |  |  |  |  | TOTAL SPAWNERS 19,20,21,23 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | River returns SFA 19-21 |  | $\begin{gathered} \hline \text { Comm- } \\ \text { ercial } \\ 19-21 \end{gathered}$ | SFA 23 |  |  |  |  | $\begin{array}{r} \text { angled } \\ 19-21 \end{array}$ | Spawners$19-21$ |  | SFA 23 |  |  |  |  |
|  |  |  | Wild | Wild | H | As 19, | ,21,23 | $\begin{aligned} & \mathrm{H}+\mathrm{W} \\ & \mathrm{MIN} \end{aligned}$ |  |  |  | rtns MAX | Harvest |  |  |
|  | MIN | MAX |  | MIN | MAX |  | MIN |  |  | MAX | MIN |  |  | MAX | MIN | MAX |
| 1970 | 8,236 | 16,868 |  | 3,189 | 5,206 | 7,421 | 100 | 16,731 | 27,578 | 3,609 | 4,627 | 13,259 | 5,306 | 7,521 | 1,420 | 8,513 | 19,360 |
| 1971 | 6,345 | 13,062 | 1,922 | 2,883 | 4,176 | 365 | 11,515 | 19,525 | 2,761 | 3,584 | 10,301 | 3,248 | 4,541 | 2,032 | 4,800 | 12,810 |
| 1972 | 6,636 | 13,354 | 1,055 | 1,546 | 2,221 | 285 | 9,522 | 16,915 | 2,917 | 3,719 | 10,437 | 1,831 | 2,506 | 2,558 | 2,992 | 10,385 |
| 1973 | 8,225 | 16,744 | 1,067 | 3,509 | 5,047 | 1,965 | 14,766 | 24,823 | 3,604 | 4,621 | 13,140 | 5,474 | 7,012 | 1,437 | 8,658 | 18,715 |
| 1974 | 14,478 | 29,385 | 2,050 | 6,204 | 8,910 | 3,991 | 26,723 | 44,336 | 6,340 | 8,138 | 23,045 | 10,195 | 12,901 | 2,124 | 16,209 | 33,822 |
| 1975 | 5,096 | 10,393 | 2,822 | 11,648 | 16,727 | 6,374 | 25,940 | 36,316 | 2,227 | 2,869 | 8,166 | 18,022 | 23,101 | 2,659 | 18,232 | 28,608 |
| 1976 | 12,421 | 25,398 | 1,675 | 13,761 | 19,790 | 9,074 | 36,931 | 55,937 | 5,404 | 7,017 | 19,994 | 22,835 | 28,864 | 5,263 | 24,589 | 43,595 |
| 1977 | 13,349 | 27,943 | 3,773 | 6,746 | 9,679 | 6,992 | 30,860 | 48,387 | 5,841 | 7,508 | 22,102 | 13,738 | 16,671 | 4,542 | 16,704 | 34,231 |
| 1978 | 2,535 | 5,241 | 3,651 | 3,227 | 4,651 | 3,044 | 12,457 | 16,587 | 1,113 | 1,422 | 4,128 | 6,271 | 7,695 | 2,015 | 5,678 | 9,808 |
| 1979 | 12,365 | 25,381 | 3,154 | 11,529 | 16,690 | 3,827 | 30,875 | 49,052 | 5,428 | 6,937 | 19,953 | 15,356 | 20,517 | 3,716 | 18,577 | 36,754 |
| 1980 | 16,534 | 33,825 | 8,252 | 14,346 | 20,690 | 10,793 | 49,925 | 73,560 | 7,253 | 9,281 | 26,572 | 25,139 | 31,483 | 5,542 | 28,878 | 52,513 |
| 1981 | 18,594 | 38,329 | 1,951 | 11,199 | 16,176 | 5,627 | 37,371 | 62,083 | 8,163 | 10,431 | 30,166 | 16,826 | 21,803 | 9,021 | 18,236 | 42,948 |
| 1982 | 10,008 | 20,552 | 2,020 | 8,773 | 12,598 | 3,038 | 23,839 | 38,208 | 4,361 | 5,647 | 16,191 | 11,811 | 15,636 | 5,279 | 12,179 | 26,548 |
| 1983 | 4,662 | 9,562 | 1,621 | 7,706 | 11,028 | 1,564 | 15,553 | 23,775 | 2,047 | 2,615 | 7,515 | 9,270 | 12,592 | 4,138 | 7,747 | 15,969 |
| 1984 | 12,398 | 25,815 | 0 | 14,105 | 20,227 | 1,451 | 27,954 | 47,493 | 4,724 | 7,674 | 21,091 | 15,556 | 21,678 | 5,266 | 17,964 | 37,503 |
| 1985 | 16,354 | 34,055 | 0 | 11,038 | 15,910 | 2,018 | 29,410 | 51,983 | 6,360 | 9,994 | 27,695 | 13,056 | 17,928 | 4,892 | 18,158 | 40,731 |
| 1986 | 16,661 | 34,495 | 0 | 13,412 | 19,321 | 862 | 30,935 | 54,678 | 6,182 | 10,479 | 28,313 | 14,274 | 20,183 | 3,549 | 21,204 | 44,947 |
| 1987 | 18,388 | 37,902 | 0 | 10,030 | 14,334 | 3,328 | 31,746 | 55,564 | 7,056 | 11,332 | 30,846 | 13,358 | 17,662 | 3,101 | 21,589 | 45,407 |
| 1988 | 16,611 | 33,851 | 0 | 15,131 | 21,834 | 1,250 | 32,992 | 56,935 | 6,384 | 10,227 | 27,467 | 16,381 | 23,084 | 3,320 | 23,288 | 47,231 |
| 1989 | 17,378 | 35,141 | 0 | 16,240 | 23,182 | 1,339 | 34,957 | 59,662 | 6,629 | 10,749 | 28,512 | 17,579 | 24,521 | 4,455 | 23,873 | 48,578 |
| 1990 | 20,119 | 41,652 | 0 | 12,287 | 17,643 | 1,533 | 33,939 | 60,828 | 7,391 | 12,728 | 34,261 | 13,820 | 19,176 | 3,795 | 22,753 | 49,642 |
| 1991 | 6,718 | 13,870 | 0 | 10,602 | 15,246 | 2,439 | 19,759 | 31,555 | 2,399 | 4,319 | 11,471 | 13,041 | 17,685 | 3,546 | 13,814 | 25,610 |
| 1992 | 9,269 | 18,936 | 0 | 11,340 | 16,181 | 2,223 | 22,832 | 37,340 | 3,629 | 5,640 | 15,307 | 13,563 | 18,404 | 4,078 | 15,125 | 29,633 |
| 1993 | 9,104 | 18,711 | 0 | 7,610 | 8,828 | foot- | 16,714 | 27,539 | 3,327 | 5,777 | 15,384 | 5,762 | 6,868 | foot- | 11,539 | 22,252 |
| 1994 | 2,446 | 4,973 | 0 | 5,770 | 6,610 | note:"a" | 8,216 | 11,583 | 493 | 1,953 | 4,480 | 4,965 | 5,738 | note:"a | 6,918 | 10,218 |
| 1995 | 5,974 | 12,364 | 0 | 8,265 | 9,458 |  | 14,239 | 21,822 | 1,885 | 4,089 | 10,479 | 8,025 | 9,218 |  | 12,114 | 19,697 |
| 1996 | 9,888 | 20,791 | 0 | 12,907 | 15,256 |  | 22,795 | 36,047 | 2,211 | 7,677 | 18,580 | 11,576 | 13,892 |  | 19,253 | 32,472 |
| 1997 | 2,665 | 5,488 | 0 | 4,508 | 4,979 |  | 7,173 | 10,467 | 493 | 2,172 | 4,995 | 3,971 | 4,433 |  | 6,143 | 9,428 |
| 1998 | 7,567 | 15,680 | 0 | 9,203 | 10,801 |  | 16,770 | 26,481 | 0 | 7,567 | 15,680 | 8,775 | 10,348 |  | 16,342 | 26,028 |
| 1999 | 5,048 | 10,535 | 0 | 5,508 | 6,366 |  | 10,556 | 16,901 | 67 | 4,981 | 10,468 | 5,196 | 6,048 |  | 10,177 | 16,516 |
| 2000 | 6,201 | 12,890 | 0 | 4,796 | 5,453 |  | 10,997 | 18,343 | 0 | 6,201 | 12,890 | 4,455 | 5,087 |  | 10,656 | 17,977 |

SFAs 19, 20, 21: Returns, 1970-1997, estimated as run size (1SW recreational catch / expl. rate [ 0.2 to 0.45 ]; where MIN and MAX selected as 5th and 95th percentile values from 1,000 monte carlo estimates) + estimated 1SW fish in commercial landings 1970-1983 (Cutting MS 1984). For 1998-2000, see "a" below.
SFA 22: Inner Fundy stocks and inner-Fundy SFA 23 (primarily 1SW fish) do not go to the North Atlantic.
SFA 23: For 1970-97, similar to SFAs 19-21 except that estimated wild 1SW returns destined for Mactaquac Dam, Saint John River, replaced values for recreational catch and estimated proportions that production above Mactaquac is of the total ( $0.4-0.6$ ) river replaced exploitataion rates (commercial
harvest, bi-catch etc., incl. in estimated returns); hatchery returns attributed to above Mactaquac only; 1SW production in rest of SFA (outer Fundy) omitted.
"a"- Revision of method, SFA 23, 1993-2000, estimated returns to Nashwaak fence raised by proportion of area below Mactaquac ( $0.21-0.30$ ) and added to total estimated returns originating upriver of Mactaquac (Marshall et al. 1998); MIN and MAX removals below Mactaquac based on Nashwaak losses, Mactaquac losses are a single value and together summed and removed from returns to establish estimate of spawners. SFAs 19-21, estimate of returns 1998-2000 based on regression of LaHave wild counts on MIN and MAX estimates of total SFA 19-21 returns, 1984-1997, because there was no (1998 and 2000) \& little (1999) angling in SFAs 20-21

Appendix 5(ixa). Total 2 SW returns to SFAs 19, 20, 21 and 23, 1970-2000.

| Year | $\begin{gathered} \text { SFA } 19 \\ \text { MIN MAX } \\ \text { SW }=0.7-0.9 \\ \text { Exp. rate }=0.2-0.45 \\ \hline \end{gathered}$ |  | $\begin{gathered} \text { SFA } 20 \\ \text { MIN MAX } \\ 2 \text { SW }=0.6-0.9 \\ \text { Exp. rate }=0.2-0.45 \\ \hline \end{gathered}$ |  | $\begin{gathered} \text { SFA } 21 \\ \text { MIN MAX } \\ 2 \text { SW }=0.5-0.9 \\ \text { Exp. rate }=0.2-0.45 \\ \hline \end{gathered}$ |  | Total Commercial 19-21 | SFA 23 |  |  |  | total returns SFAs 19,20,21,23 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | W ild W ild <br> M IN MAX <br> $2 S W=0.85-0.95$  <br> p. $\mathrm{abv}=0.4-0.6$  |  |  |  | Htch Htch <br> MIN MAX <br> $2 S W=0.85-0.95$  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  | M IN | M AX |  |  |
| 1970 | 1,170 | 2,537 |  |  | 658 | 1,535 | 597 | 1,525 | 2,644 | 8,540 | 12,674 | 0 | 0 | 13,609 | 20,915 |
| 1971 | 600 | 1,266 | 344 | 802 | 481 | 1,199 | 2,607 | 7,089 | 10,463 | 66 | 73 | 11,187 | 16,410 |
| 1972 | 735 | 1,614 | 421 | 1,002 | 454 | 1,198 | 4,549 | 7,362 | 10,809 | 507 | 559 | 14,028 | 19,731 |
| 1973 | 726 | 1,571 | 665 | 1,532 | 546 | 1,437 | 4,217 | 3,773 | 5,559 | 432 | 477 | 10,359 | 14,793 |
| 1974 | 1,035 | 2,225 | 691 | 1,588 | 548 | 1,397 | 8,873 | 8,766 | 12,790 | 1,989 | 2,198 | 21,902 | 29,071 |
| 1975 | 376 | 824 | 149 | 343 | 882 | 2,321 | 9,430 | 11,217 | 16,490 | 1,890 | 2,088 | 23,944 | 31,496 |
| 1976 | 791 | 1,672 | 346 | 822 | 441 | 1,146 | 5,916 | 12,304 | 18,106 | 1,970 | 2,175 | 21,768 | 29,837 |
| 1977 | 999 | 2,152 | 660 | 1,509 | 873 | 2,354 | 9,205 | 14,539 | 21,420 | 2,330 | 2,575 | 28,606 | 39,215 |
| 1978 | 810 | 1,739 | 429 | 995 | 655 | 1,706 | 6,827 | 6,059 | 8,903 | 2,166 | 2,391 | 16,946 | 22,561 |
| 1979 | 532 | 1,169 | 431 | 978 | 508 | 1,288 | 2,326 | 4,149 | 6,084 | 1,016 | 1,123 | 8,962 | 12,968 |
| 1980 | 1,408 | 3,051 | 746 | 1,714 | 1,483 | 3,989 | 9,204 | 16,500 | 24,041 | 2,556 | 2,824 | 31,897 | 44,823 |
| 1981 | 886 | 1,856 | 926 | 2,133 | 1,754 | 4,475 | 4,438 | 8,696 | 12,690 | 2,330 | 2,577 | 19,030 | 28,169 |
| 1982 | 917 | 1,990 | 316 | 746 | 682 | 1,756 | 5,819 | 8,266 | 12,198 | 1,516 | 1,673 | 17,516 | 24,182 |
| 1983 | 477 | 1,030 | 641 | 1,475 | 552 | 1,434 | 2,978 | 8,718 | 12,793 | 944 | 1,043 | 14,310 | 20,753 |
| 1984 | 828 | 1,768 | 638 | 1,500 | 766 | 2,004 | 0 | 14,753 | 21,573 | 953 | 1,054 | 17,938 | 27,899 |
| 1985 | 1,495 | 3,132 | 2,703 | 6,355 | 2,102 | 5,469 | 0 | 15,793 | 23,002 | 748 | 826 | 22,841 | 38,784 |
| 1986 | 3,500 | 7,541 | 2,561 | 5,987 | 2,150 | 5,312 | 0 | 9,210 | 13,507 | 681 | 754 | 18,102 | 33,101 |
| 1987 | 2,427 | 5,237 | 1,066 | 2,527 | 1,114 | 2,872 | 0 | 6,512 | 9,590 | 410 | 453 | 11,529 | 20,679 |
| 1988 | 2,635 | 5,724 | 1,914 | 4,464 | 1,105 | 2,945 | 0 | 3,936 | 5,836 | 780 | 861 | 10,370 | 19,830 |
| 1989 | 2,236 | 4,810 | 1,512 | 3,485 | 1,631 | 4,086 | 0 | 6,159 | 8,994 | 401 | 443 | 11,939 | 21,818 |
| 1990 | 2,406 | 5,178 | 1,085 | 2,515 | 1,271 | 3,260 | 0 | 4,994 | 7,375 | 492 | 543 | 10,248 | 18,871 |
| 1991 | 1,890 | 4,050 | 965 | 2,200 | 421 | 1,071 | 0 | 6,739 | 9,902 | 598 | 661 | 10,613 | 17,884 |
| 1992 | 1,788 | 3,923 | 631 | 1,488 | 480 | 1,236 | 0 | 6,213 | 9,074 | 665 | 735 | 9,777 | 16,456 |
| 1993 | 876 | 1,897 | 1,006 | 2,321 | 564 | 1,498 | 0 | 4,318 | 5,371 |  |  | 6,764 | 11,087 |
| 1994 | 833 | 1,845 | 242 | 561 | 305 | 773 | 0 | 2,999 | 3,729 | note |  | 4,379 | 6,908 |
| 1995 | 759 | 1,582 | 666 | 1,565 | 518 | 1,339 | 0 | 3,042 | 3,831 |  |  | 4,985 | 8,317 |
| 1996 | 1,231 | 2,692 | 604 | 1,404 | 894 | 2,293 | 0 | 4,498 | 5,665 |  |  | 7,227 | 12,054 |
| 1997 | 607 | 1,299 | 170 | 387 | 301 | 1,026 | 0 | 2,567 | 3,210 |  |  | 3,645 | 5,922 |
| 1998 | >>>>>>>> | >>>>>> | ->>>>>>>> | >>>>> | 1,103 | 3,888 | 0 | 1,625 | 2,115 |  |  | 2,728 | 6,003 |
| 1999 | >>> | - | ->>>>>> | $>$ | 1,230 | 4,324 | 0 | 2,252 | 2,783 |  |  | 3,482 | 7,107 |
| 2000 | >>>>>>>> | >>>> | >>>>>>> | >>>>> | 1,086 | 3,816 | , | 952 | 1,263 |  |  | 2,038 | 5,079 |

SFAs 19, 20, 21: Returns, 1970-97 estimated as run size (MSW recreational catch *prop. 2SW [range of values]/expl. rate [range of values]; where MIN and MAX selected as 5 th and 95th percentile values from 1,000 monte carlo estimates) + estimated 2 SW fish in commercial landings 1970-1983 (Cutting MS 1984). For 1998-2000 see "a" below.
$\qquad$ Inner Fundy stocks do not go to north Atlantic.
SFA 23: For 1970-1997 Similar approach as for SFAs 19-21 except thatestimated wild MSW returns destined for Mactaquac Dam, Saint John River, replaced values fo recreational catch; and estimated proportions that production above Mactaquac is of the total river replaced exploitation rates (commercial harvest, bi-catch etc., incl. in estimated returns) + est. $0.85-0.95^{*} \mathrm{MSW}$ hatchery returns to Mactaquac; 2 SW production in rest of SFA omitted.
"a": Revsion of method, SFA 23, 1993-2000, estimated MSW returns to Nashwaak fence raised by prop. of area below Mactaquac ( $0.21-0.30$ ) * prop. 2SW ( 0.7 \& 0.9 ) and added to estimated MSW hatchery and wild returns * (Marshall et al. MS 1998) (0.85-0.95; 2SW) originating upriver of Mactaquac. MIN \& MAX removals below Mactaquac based on Nashwaak losses: Mactaquac losses were a single value and together summed and removed from MSW returns (prevously) to estimate spawners. SFAs 19-21, estimate of 2 SW returns for 1998-00, based on regression of LaHave wild counts on MIN and MAX estimates of total SFA 19-21 MSW returns and 5 th and 95 th percentile values of MIN-MAX ( $0.5 \& 0.92$ SW fish among MSW salmon).

Appendix 5(ixb). Total 2SW spawners in SFAs 19, 20, 21 and 23, 1970-2000.

| Year | SFA 19 |  | RETURNS SFA 20 |  | SFA 21 |  | REMOVALS angled (19-21) |  | SPAWNERS <br> SFAs (19-21) |  | SFA 23 |  |  |  | TOTAL SPAWNERS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | RETURNS | REMOVALS |  |  |  |  |  |  |  |
|  | MIN | MAX |  |  | MIN | MAX | MIN | MAX | MIN | MAX | MIN | MAX | MIN | MAX | MIN | MAX | MIN | MAX |
| 1970 | 1,170 | 2,537 | 658 | 1,535 |  |  | 597 | 1,525 | 941 | 1,375 | 1,485 | 4,222 | 8,540 | 12,674 | 7,004 | 7,828 | 3,021 | 9,068 |
| 1971 | 600 | 1,266 | 344 | 802 | 481 | 1,199 | 541 | 812 | 884 | 2,455 | 7,155 | 10,536 | 3,543 | 3,960 | 4,496 | 9,032 |
| 1972 | 735 | 1,614 | 421 | 1,002 | 454 | 1,198 | 623 | 922 | 987 | 2,892 | 7,869 | 11,368 | 1,397 | 1,562 | 7,459 | 12,699 |
| 1973 | 726 | 1,571 | 665 | 1,532 | 546 | 1,437 | 740 | 1,108 | 1,197 | 3,432 | 4,205 | 6,036 | 1,454 | 1,625 | 3,949 | 7,844 |
| 1974 | 1,035 | 2,225 | 691 | 1,588 | 548 | 1,397 | 871 | 1,277 | 1,404 | 3,933 | 10,755 | 14,988 | 2,632 | 2,942 | 9,526 | 15,979 |
| 1975 | 376 | 824 | 149 | 343 | 882 | 2,321 | 534 | 867 | 874 | 2,621 | 13,107 | 18,578 | 2,120 | 2,369 | 11,861 | 18,830 |
| 1976 | 791 | 1,672 | 346 | 822 | 441 | 1,146 | 603 | 887 | 975 | 2,754 | 14,274 | 20,281 | 4,203 | 4,698 | 11,045 | 18,337 |
| 1977 | 999 | 2,152 | 660 | 1,509 | 873 | 2,354 | 967 | 1,463 | 1,565 | 4,552 | 16,869 | 23,995 | 4,856 | 5,427 | 13,578 | 23,119 |
| 1978 | 810 | 1,739 | 429 | 995 | 655 | 1,706 | 723 | 1,088 | 1,171 | 3,352 | 8,225 | 11,294 | 2,879 | 3,218 | 6,517 | 11,428 |
| 1979 | 532 | 1,169 | 431 | 978 | 508 | 1,288 | 560 | 851 | 911 | 2,585 | 5,165 | 7,207 | 1,393 | 1,557 | 4,683 | 8,234 |
| 1980 | 1,408 | 3,051 | 746 | 1,714 | 1,483 | 3,989 | 1,390 | 2,131 | 2,247 | 6,623 | 19,056 | 26,865 | 7,033 | 7,860 | 14,270 | 25,628 |
| 1981 | 886 | 1,856 | 926 | 2,133 | 1,754 | 4,475 | 1,338 | 2,125 | 2,228 | 6,339 | 11,026 | 15,267 | 7,384 | 8,253 | 5,870 | 13,353 |
| 1982 | 917 | 1,990 | 316 | 746 | 682 | 1,756 | 734 | 1,096 | 1,181 | 3,396 | 9,782 | 13,871 | 5,307 | 5,932 | 5,656 | 11,335 |
| 1983 | 477 | 1,030 | 641 | 1,475 | 552 | 1,434 | 633 | 971 | 1,037 | 2,968 | 9,662 | 13,836 | 9,194 | 10,275 | 1,505 | 6,529 |
| 1984 | 828 | 1,768 | 638 | 1,500 | 766 | 2,004 | 267 | 419 | 1,965 | 4,853 | 15,706 | 22,627 | 3,426 | 3,829 | 14,245 | 23,650 |
| 1985 | 1,495 | 3,132 | 2,703 | 6,355 | 2,102 | 5,469 |  |  | 6,300 | 14,956 | 16,541 | 23,828 | 4,656 | 5,204 | 18,185 | 33,580 |
| 1986 | 3,500 | 7,541 | 2,561 | 5,987 | 2,150 | 5,312 |  |  | 8,211 | 18,840 | 9,891 | 14,261 | 2,667 | 2,981 | 15,435 | 30,120 |
| 1987 | 2,427 | 5,237 | 1,066 | 2,527 | 1,114 | 2,872 |  |  | 4,607 | 10,636 | 6,922 | 10,043 | 1,294 | 1,446 | 10,235 | 19,233 |
| 1988 | 2,635 | 5,724 | 1,914 | 4,464 | 1,105 | 2,945 |  |  | 5,654 | 13,133 | 4,716 | 6,697 | 1,296 | 1,449 | 9,074 | 18,381 |
| 1989 | 2,236 | 4,810 | 1,512 | 3,485 | 1,631 | 4,086 |  |  | 5,379 | 12,381 | 6,560 | 9,437 | 250 | 279 | 11,689 | 21,539 |
| 1990 | 2,406 | 5,178 | 1,085 | 2,515 | 1,271 | 3,260 |  |  | 4,762 | 10,953 | 5,486 | 7,918 | 560 | 626 | 9,688 | 18,245 |
| 1991 | 1,890 | 4,050 | 965 | 2,200 | 421 | 1,071 |  |  | 3,276 | 7,321 | 7,337 | 10,563 | 1,257 | 1,405 | 9,356 | 16,479 |
| 1992 | 1,788 | 3,923 | 631 | 1,488 | 480 | 1,236 |  |  | 2,899 | 6,647 | 6,878 | 9,809 | 1,052 | 1,176 | 8,725 | 15,280 |
| 1993 | 876 | 1,897 | 1,006 | 2,321 | 564 | 1,498 |  |  | 2,446 | 5,716 | 4,318 | 5,371 | 1,054 | 1,166 | 5,710 | 9,921 |
| 1994 | 833 | 1,845 | 242 | 561 | 305 | 773 |  |  | 1,380 | 3,179 | 2,999 | 3,729 | 697 | 815 | 3,682 | 6,093 |
| 1995 | 759 | 1,582 | 666 | 1,565 | 518 | 1,339 |  |  | 1,943 | 4,486 | 3,042 | 3,831 | 313 | 346 | 4,672 | 7,971 |
| 1996 | 1,231 | 2,692 | 604 | 1,404 | 894 | 2,293 |  |  | 2,729 | 6,389 | 4,498 | 5,665 | 720 | 812 | 6,507 | 11,242 |
| 1997 | 607 | 1,299 | 170 | 387 | 301 | 1,026 |  |  | 1,078 | 2,712 | 2,567 | 3,210 | 550 | 611 | 3,095 | 5,311 |
| 1998 | >>>>>>>> | >>>>> | ->>>> | >>>> | 1,103 | 3,888 |  |  | 1,103 | 3,888 | 1,625 | 2,115 | 304 | 340 | 2,424 | 5,663 |
| 1999 | ->> | ->>>> | - | - | 1,230 | 4,324 |  |  | 1,230 | 4,324 | 2,252 | 2,783 | 441 | 459 | 3,041 | 6,648 |
| 2000 | >>>>>>>> |  | ->>>> | - | 1,086 | 3,816 |  |  | 1,086 | 3,816 | 952 | 1,263 | 183 | 202 | 1,855 | 4,877 |

Spawners = returns minus removals where "returns" are from previous Appendix as are outlines of revisions to methods for SFAs 19-21, 1998-2000, and SFA 23, 1993-2000.
"Removals" of 2SW fish in SFAs 19-21 have been few, largely illegal and unascribed since the catch-and-release angling regulations in 1985; removals in SFA 23, 1985-1997,
had been in total, the assessed losses to stocks originating above Mactaquac. The revised method, 1993-2000, incorporates 5th and 95th percentile values for losses
noted on the Nashwaak raised to the total production area downstream of Mactaquac as well as the previously assessed and used values for stocks upstream of Mactaquac.

Appendix 5(x). Estimated numbers of salmons recruits and spawners for Québec, 1969-2000.

| Year | Recruit of small salmon |  |  | Recruit of large salmon |  |  | Spawner of small salmon |  |  | Spawner of large salmon |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Min | Mean | Max | Min | Mean | Max | Min | Mean | Max | Min | Mean | Max |
| 1969 | 25355 | 31694 | 38032 | 74653 | 93316 | 111979 | 16313 | 20392 | 24470 | 25532 | 31915 | 38299 |
| 1970 | 18904 | 23630 | 28356 | 82680 | 103350 | 124020 | 11045 | 13806 | 16568 | 31292 | 39115 | 46937 |
| 1971 | 14969 | 18711 | 22453 | 47354 | 59192 | 71031 | 9338 | 11672 | 14007 | 16194 | 20243 | 24292 |
| 1972 | 12470 | 15587 | 18704 | 61773 | 77217 | 92660 | 8213 | 10267 | 12320 | 31727 | 39658 | 47590 |
| 1973 | 16585 | 20731 | 24877 | 68171 | 85214 | 102256 | 10987 | 13734 | 16480 | 32279 | 40349 | 48419 |
| 1974 | 16791 | 20988 | 25186 | 91455 | 114319 | 137182 | 10067 | 12583 | 15100 | 39256 | 49070 | 58884 |
| 1975 | 18071 | 22589 | 27106 | 77664 | 97080 | 116497 | 11606 | 14507 | 17409 | 32627 | 40784 | 48940 |
| 1976 | 19959 | 24948 | 29938 | 77212 | 96515 | 115818 | 12979 | 16224 | 19469 | 31032 | 38790 | 46548 |
| 1977 | 18190 | 22737 | 27285 | 91017 | 113771 | 136525 | 12004 | 15005 | 18006 | 44660 | 55825 | 66990 |
| 1978 | 16971 | 21214 | 25456 | 81953 | 102441 | 122930 | 11447 | 14309 | 17170 | 40944 | 51180 | 61416 |
| 1979 | 21683 | 27103 | 32524 | 45197 | 56497 | 67796 | 15863 | 19829 | 23795 | 17543 | 21929 | 26315 |
| 1980 | 29791 | 37239 | 44686 | 107461 | 134327 | 161192 | 20817 | 26021 | 31226 | 48758 | 60948 | 73137 |
| 1981 | 41667 | 52084 | 62501 | 84428 | 105535 | 126642 | 30952 | 38690 | 46428 | 35798 | 44747 | 53697 |
| 1982 | 23699 | 29624 | 35549 | 74870 | 93587 | 112305 | 16877 | 21096 | 25316 | 36290 | 45363 | 54435 |
| 1983 | 17987 | 22484 | 26981 | 61488 | 76860 | 92232 | 12030 | 15038 | 18045 | 23710 | 29638 | 35565 |
| 1984 | 21566 | 26230 | 30894 | 61180 | 71110 | 81041 | 16316 | 20636 | 24957 | 30610 | 37674 | 44739 |
| 1985 | 22771 | 28016 | 33262 | 62899 | 73545 | 84192 | 15608 | 20374 | 25140 | 28312 | 35897 | 43482 |
| 1986 | 33758 | 40347 | 46937 | 75561 | 87479 | 99397 | 22230 | 28042 | 33855 | 32997 | 41114 | 49232 |
| 1987 | 37816 | 45925 | 54034 | 72190 | 82920 | 93650 | 25789 | 33135 | 40481 | 29758 | 36610 | 43462 |
| 1988 | 43943 | 53068 | 62193 | 77904 | 90587 | 103269 | 28582 | 36699 | 44815 | 34781 | 43653 | 52524 |
| 1989 | 34568 | 41488 | 48407 | 70762 | 81316 | 91871 | 24710 | 31015 | 37319 | 34268 | 41727 | 49185 |
| 1990 | 39962 | 47377 | 54792 | 68851 | 79872 | 90893 | 26594 | 33210 | 39826 | 33454 | 41535 | 49615 |
| 1991 | 31488 | 37121 | 42755 | 64166 | 73675 | 83184 | 20582 | 25508 | 30433 | 27341 | 33569 | 39797 |
| 1992 | 35257 | 42000 | 48742 | 64271 | 74112 | 83953 | 21754 | 27668 | 33583 | 26489 | 32993 | 39497 |
| 1993 | 30645 | 36400 | 42156 | 50717 | 57197 | 63677 | 17493 | 22469 | 27444 | 21609 | 25481 | 29353 |
| 1994 | 29667 | 34918 | 40170 | 51649 | 58139 | 64630 | 16758 | 21200 | 25642 | 21413 | 25191 | 28968 |
| 1995 | 23851 | 28109 | 32368 | 59939 | 67083 | 74227 | 14409 | 17978 | 21548 | 30925 | 35122 | 39320 |
| 1996 | 32008 | 37283 | 42558 | 53990 | 61136 | 68282 | 18923 | 23364 | 27805 | 26042 | 30433 | 34824 |
| 1997 | 24300 | 28659 | 33018 | 44442 | 50315 | 56187 | 14724 | 18467 | 22210 | 21275 | 24871 | 28466 |
| 1998 | 24029 | 28777 | 33524 | 33280 | 38370 | 43460 | 16277 | 20615 | 24954 | 19419 | 22951 | 26483 |
| 1999 | 25639 | 30967 | 36296 | 34742 | 40400 | 46058 | 18785 | 23643 | 28502 | 23559 | 28028 | 32497 |
| 2000 | 25216 | 30762 | 36309 | 33829 | 40574 | 47320 | 17137 | 22116 | 27096 | 22009 | 27008 | 32007 |
| Mean 95-99 | 25965 | 30759 | 35553 | 45279 | 51461 | 57643 | 16624 | 20814 | 25004 | 24244 | 28281 | 32318 |
| 2000 vs |  |  |  |  |  |  |  |  |  |  |  |  |
| 1999 |  | -1 |  |  | 0 |  |  | -6 |  |  | -4 |  |
| 1995-1999 |  | 0 |  |  | -21 |  |  | 6 |  |  | -5 |  |

## APPENDIX 6

## Computation of Catch Advice for West Greenland

The North American Spawning Reserve (SpT) for 2 SW salmon of 152,548 fish remains the same as in 2000.

This number must be divided by the survival rate for the fish from the time of the West Greenland fishery to their return of the fish to home waters (11 months) to give the Spawning Target Reserve (SpR). Thus:

Eq. 1. $\quad \mathrm{SpR}=\mathrm{SpT} *(\exp (11 * \mathrm{M}) \quad($ where $\mathrm{M}=0.01)$

The Maximum Allowable Harvest (MAH) may be defined as the number of non-maturing 1SW fish that are available for harvest. This number is calculated by subtracting the Spawning Target Reserve from the pre-fishery abundance (PFA).

Eq. 2. $\quad \mathrm{MAH}=\mathrm{PFA}-\mathrm{SpR}$
To provide catch advice for West Greenland it is then necessary to decide on the proportion of the MAH to be allocated to Greenland ( $\mathrm{f}_{\mathrm{NA}}$ ). The allowable harvest of North American non-maturing 1SW salmon at West Greenland NA1SW) may then be defined as

Eq. 3. $\quad \mathrm{NA} 1 \mathrm{SW}=\mathrm{f}_{\mathrm{NA}} * \mathrm{MAH}$
The estimated number of European salmon that will be caught at West Greenland (E1SW) will depend upon the harvest of North American fish and the proportion of the fish in the West Greenland fishery that originate from North America [PropNA] ${ }^{1}$. Thus:

Eq. 4. $\quad E 1 S W=(N A 1 S W /$ PropNA $)-$ NA1SW

To convert the numbers of North American and European 1SW salmon into total catch at West Greenland in metric tonnes, it is necessary to incorporate the mean weights (kg) of salmon for North America [WT1SWNA] ${ }^{1}$ and Europe [WT1SWE] ${ }^{1}$ and age correction factor for multi-sea winter salmon at Greenland based on the total weight of salmon caught divided by the weight of 1 SW salmon $[\mathrm{ACF}]^{1}$. The quota (in tonnes) at Greenland is then estimated as

Eq. 5. $\quad$ Quota $=($ NA1SW $* W T 1 S W N A+E 1 S W * W T 1 S W E) * A C F / 1000$
1 Sampling data from the 1995-99 fishery at West Greenland were used to update the forecast values by exponential smoothing of the proportion of North American salmon in the catch (PropNA), weights by continent [WT1SWNA, WT1SWE] and the age correction factor [ACF].

## APPENDIX 7

Appendix 7a Input data for NEAC Area Pre Fishery Abundance analysis using Monte Carlo simulation - FINLAND

| Year | Catch (numbers) 1SW | MSW | Unrep. as \% of total 1SW <br> min | $\max$ | Unrep. as \% of total MSW $\qquad$ | max | $\begin{array}{\|l} \text { Exp. } \\ \text { rate } \\ 1 S W \\ (\%) \\ (\%) \\ \text { min } \end{array}$ | $\max$ | Exp. rate MSW (\%) <br> min | max |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 3,114 | 3,156 | 20 | 30 | 20 | 30 | 40 | 70 | 40 | 80 |
| 1972 | 4,865 | 4,932 | 20 | 30 | 20 | 30 | 40 | 70 | 40 | 80 |
| 1973 | 7,395 | 7,496 | 20 | 30 | 20 | 30 | 40 | 70 | 40 | 80 |
| 1974 | 6,803 | 7,253 | 20 | 30 | 20 | 30 | 40 | 70 | 40 | 80 |
| 1975 | 6,732 | 7,178 | 20 | 30 | 20 | 30 | 40 | 70 | 40 | 80 |
| 1976 | 5,817 | 6,202 | 20 | 30 | 20 | 30 | 40 | 70 | 40 | 80 |
| 1977 | 5,238 | 5,584 | 20 | 30 | 20 | 30 | 40 | 70 | 40 | 80 |
| 1978 | 3,832 | 3,481 | 20 | 30 | 20 | 30 | 40 | 70 | 40 | 80 |
| 1979 | 3,982 | 2,298 | 20 | 30 | 20 | 30 | 40 | 70 | 30 | 70 |
| 1980 | 3,920 | 3,093 | 20 | 30 | 20 | 30 | 40 | 70 | 30 | 70 |
| 1981 | 3,617 | 4,874 | 20 | 30 | 20 | 30 | 40 | 70 | 30 | 70 |
| 1982 | 2,598 | 5,408 | 20 | 30 | 20 | 30 | 40 | 70 | 30 | 70 |
| 1983 | 3,916 | 6,050 | 20 | 30 | 20 | 30 | 40 | 70 | 30 | 70 |
| 1984 | 4,899 | 4,726 | 20 | 30 | 20 | 30 | 40 | 70 | 30 | 70 |
| 1985 | 6,201 | 4,912 | 20 | 30 | 20 | 30 | 40 | 70 | 30 | 70 |
| 1986 | 6,131 | 3,244 | 20 | 30 | 20 | 30 | 40 | 70 | 30 | 70 |
| 1987 | 8,696 | 4,520 | 20 | 30 | 20 | 30 | 40 | 70 | 30 | 70 |
| 1988 | 5,926 | 3,495 | 20 | 30 | 20 | 30 | 40 | 70 | 30 | 70 |
| 1989 | 10,395 | 5,332 | 20 | 30 | 20 | 30 | 50 | 80 | 40 | 80 |
| 1990 | 10,084 | 5,600 | 20 | 30 | 20 | 30 | 50 | 80 | 40 | 80 |
| 1991 | 9,213 | 6,298 | 20 | 30 | 20 | 30 | 50 | 80 | 40 | 80 |
| 1992 | 15,017 | 6,284 | 20 | 30 | 20 | 30 | 50 | 80 | 40 | 80 |
| 1993 | 11,157 | 8,180 | 20 | 30 | 20 | 30 | 50 | 80 | 40 | 80 |
| 1994 | 7,493 | 6,230 | 20 | 30 | 20 | 30 | 50 | 80 | 40 | 80 |
| 1995 | 7,786 | 5,344 | 20 | 30 | 20 | 30 | 50 | 80 | 40 | 80 |
| 1996 | 10,726 | 2,717 | 20 | 30 | 20 | 30 | 40 | 70 | 30 | 70 |
| 1997 | 9,469 | 4,272 | 20 | 30 | 20 | 30 | 40 | 70 | 30 | 70 |
| 1998 | 11,410 | 3,749 | 20 | 30 | 20 | 30 | 40 | 70 | 30 | 70 |
| 1999 | 16,861 | 3,848 | 20 | 30 | 20 | 30 | 50 | 80 | 40 | 70 |
| 2000 | 17,499 | 8,698 | 20 | 30 | 20 | 30 | 50 | 80 | 40 | 70 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| $M(\min )=$ | 0.005 | Return time $(\mathrm{m})=1 \mathrm{SW}(\min )$ | 7 | $\mathrm{MSW}(\min )$ | 16 |
| ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathrm{M}(\max )=$ | 0.015 | $1 \mathrm{SW}(\max )$ | 9 | $\mathrm{NSW}(\max )$ | 18 |

Appendix 7b Input data for NEAC Area Pre Fishery Abundance analysis using Monte Carlo simulation - FRANCE

| Year | Catch (numbers)1SW |  | Unrep. as \% of total 1SW$\begin{array}{l\|l} \min & \max \\ \hline \end{array}$ |  | Unrep. as \% of total MSW$\begin{array}{l\|l} \mathrm{m} \text { in } & \mathrm{max} \\ \hline \end{array}$ |  | $\begin{aligned} & \text { Exp. rate } \\ & 1 \mathrm{SW}(\%) \\ & \min \\ & \hline \end{aligned}$ |  | Exp. rate MSW (\%)$\min \quad \max$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Non-reporting included in exploitation rates |  |  |  |  |  |  |  |  |  |
| 1971 | 1,740 | 4,060 | 0 | 0 | 0 | 0 | 2 | 5 | 25 | 50 |
| 1972 | 3.480 | 8,120 | 0 | 0 | 0 | 0 | 2 | 5 | 25 | 50 |
| 1973 | 2,130 | 4,970 | 0 | 0 | 0 | 0 | 2 | 5 | 25 | 50 |
| 1974 | 990 | 2,310 | 0 | 0 | 0 | 0 | 2 | 5 | 25 | 50 |
| 1975 | 1,980 | 4,620 | 0 | 0 | 0 | 0 | 2 | 5 | 25 | 50 |
| 1976 | 1,820 | 3,380 | 0 | 0 | 0 | 0 | 2 | 5 | 25 | 50 |
| 1977 | 1,400 | 2,600 | 0 | 0 | 0 | 0 | 2 | 5 | 25 | 50 |
| 1978 | 1,435 | 2,665 | 0 | 0 | 0 | 0 | 2 | 5 | 25 | 50 |
| 1979 | 1,645 | 3,055 | 0 | 0 | 0 | 0 | 2 | 5 | 25 | 50 |
| 1980 | 3.430 | 6,370 | 0 | 0 | 0 | 0 | 2 | 5 | 25 | 50 |
| 1981 | 2,720 | 4,080 | 0 | 0 | 0 | 0 | 2 | 5 | 20 | 50 |
| 1982 | 1,680 | 2,520 | 0 | 0 | 0 | 0 | 2 | 5 | 20 | 50 |
| 1983 | 1,800 | 2,700 | 0 | 0 | 0 | 0 | 2 | 5 | 20 | 50 |
| 1984 | 2,960 | 4,440 | 0 | 0 | 0 | 0 | 2 | 5 | 20 | 50 |
| 1985 | 1,100 | 3,330 | 0 | 0 | 0 | 0 | 2 | 5 | 20 | 50 |
| 1986 | 3,400 | 3,400 | 0 | 0 | 0 | 0 | 2 | 12 | 20 | 50 |
| 1987 | 6.000 | 1,800 | 0 | 0 | 0 | 0 | 2 | 12 | 20 | 50 |
| 1988 | 2,100 | 5,000 | 0 | 0 | 0 | 0 | 2 | 12 | 20 | 50 |
| 1989 | 1,100 | 2,300 | 0 | 0 | 0 | 0 | 2 | 12 | 20 | 50 |
| 1990 | 1,900 | 2,300 | 0 | 0 | 0 | 0 | 2 | 12 | 20 | 50 |
| 1991 | 1,400 | 2,100 | 0 | 0 | 0 | 0 | 2 | 12 | 20 | 50 |
| 1992 | 2,500 | 2,700 | 0 | 0 | 0 | 0 | 2 | 12 | 20 | 50 |
| 1993 | 3,600 | 1,300 | 0 | 0 | 0 | 0 | 2 | 12 | 20 | 50 |
| 1994 | 2,800 | 2,300 | 0 | 0 | 0 | 0 | 2 | 12 | 20 | 40 |
| 1995 | 1,669 | 1,095 | 0 | 0 | 0 | 0 | 5 | 20 | 20 | 40 |
| 1996 | 2,063 | 1,942 | 0 | 0 | 0 | 0 | 5 | 20 | 20 | 40 |
| 1997 | 1,060 | 1,001 | 0 | 0 | 0 | 0 | 5 | 20 | 20 | 40 |
| 1998 | 2,065 | 846 | 0 | 0 | 0 | 0 | 5 | 20 | 20 | 40 |
| 1999 | 690 | 1,831 | 0 | 0 | 0 | 0 | 5 | 20 | 20 | 40 |
| 2000 | 1,792 | 1,277 | 0 | 0 | 0 | 0 | 5 | 20 | 20 | 40 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| $M(\min )=$ | 0.005 | Return time $(\mathrm{m})=\mathrm{W}(\min )$ | 7 | $\mathrm{~W}(\min )$ | 16 |
| ---: | ---: | ---: | ---: | ---: | ---: |
| $M(\max )$ | $=0.015$ | $1 S W(\max )$ | 9 | $\mathrm{~V}(\max )$ | 18 |

Appendix 7c Input data for NEAC Area Pre Fishery Abundance analysis using Monte Carlo simulation - ICELAND-1

| Year | Catch (numbers) |  | Unrep. as \% of total 1SW |  |  |  | $\begin{aligned} & \text { Exp. rate } \\ & \text { 1SW (\%) } \end{aligned}$ |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | min | max | min | max | min | max | min | max |
| 1971 | 25,746 | 15,429 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1972 | 26,793 | 19,715 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1973 | 26,105 | 18,847 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1974 | 17,742 | 17,364 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1975 | 24,541 | 19,695 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1976 | 20,046 | 17,588 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1977 | 21,394 | 17,636 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1978 | 28,436 | 21,858 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1979 | 24,508 | 14,099 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1980 | 15,447 | 15,784 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1981 | 16,666 | 10,785 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1982 | 12.398 | 10,872 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1983 | 20,397 | 15,348 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1984 | 14,129 | 11,573 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1985 | 17,595 | 9,900 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1986 | 26,753 | 13,310 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1987 | 20,697 | 8.366 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1988 | 39,891 | 7.534 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1989 | 22.050 | 7.168 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1990 | 21,396 | 7.290 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1991 | 21,438 | 7.206 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1992 | 24,245 | 9.330 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1993 | 23,818 | 6.040 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1994 | 20,668 | 8.079 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1995 | 25,517 | 7.122 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1996 | 21,264 | 5,605 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1997 | 15,470 | 5.565 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1998 | 21,520 | 3,707 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1999 | 16,622 | 5.757 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 2000 | 11,894 | 2,358 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| $M($ min $)=$ | 0.005 |  | Retu | time ( m ) $=$ | 1SW(min) | 7 | MSW(min) | 16 |  |  |
| $\mathrm{M}($ max $)=$ | 0.015 |  |  |  | 1SW(max) | 9 | MSW(max) | 18 |  |  |

Appendix 7d Input data for NEAC Area Pre Fishery Abundance analysis using Monte Carlo simulation - ICELAND-2

| Year | Catch (numbers) 1SW | MSW | Unrep. as \% of total 1SW | max | Unrep. as $\%$ of total MSW | max | Exp. rate $1 S W$ (\%) min | max | Exp. rate MSW (\% <br> $\min$ | max |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 4,704 | 6,191 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1972 | 4,334 | 9,728 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1973 | 5,803 | 12,059 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1974 | 4.867 | 8.475 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1975 | 6,343 | 9,907 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1976 | 5,880 | 7,883 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1977 | 9,285 | 11,824 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1978 | 9,232 | 15,323 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1979 | 8,807 | 9,952 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1980 | 1,586 | 14,140 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1981 | 7,105 | 4,965 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1982 | 3,358 | 5,956 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1983 | 4,583 | 4,259 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1984 | 1.742 | 5.460 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1985 | 10,233 | 3,398 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1986 | 13,745 | 10,189 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1987 | 8,238 | 9,642 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1988 | 12,776 | 6,519 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1989 | 6,644 | 5,582 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1990 | 5,403 | 5,866 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1991 | 8,583 | 4,552 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1992 | 13,831 | 6,440 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1993 | 11,354 | 6,922 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1994 | 3,712 | 5,828 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1995 | 9,494 | 3,714 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1996 | 5,007 | 4,865 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1997 | 7.354 | 2.955 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1998 | 12,822 | 4,261 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 1999 | 5,905 | 7,077 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 2000 | 6081 | 2624 | 1 | 3 | 1 | 3 | 40 | 60 | 50 | 70 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

$\begin{array}{cc}\mathrm{M}(\min )= & 0.005 \\ \mathrm{M}(\max )= & 0.015\end{array}$

Return time $(\mathrm{m})=$
1SW(max
$\begin{array}{ll}\text { MSW(max) } & 18\end{array}$

Appendix 7e Input data for NEAC Area Pre Fishery Abundance analysis using Mbnte Carlo simulation - Al IREAND.

| Year | Catch (numbers) 1SW | MSW | Unrep. as \%of total 1SW <br> min | max | Unrep. as \%of total MSW min | max | Exp. rate 1SW (\%) <br> min | max | Exp. rate MSW (\%) <br> min | max |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 475,839 | 52871 | 30.00 | 45.00 | 30.00 | 50.00 | 66.00 | 78.00 | 47.00 | 55.00 |
| 1972 | 523,742 | 58,194 | 30.00 | 45.00 | 30.00 | 50.00 | 66.30 | 78.00 | 46.75 | 55.00 |
| 1973 | 560,323 | 62,258 | 30.00 | 45.00 | 30.00 | 50.00 | 66.30 | 78.00 | 46.75 | 55.00 |
| 1974 | 617,806 | 68,645 | 30.00 | 45.00 | 30.00 | 50.00 | 66.30 | 78.00 | 46.75 | 55.00 |
| 1975 | 643,355 | 71,484 | 30.00 | 45.00 | 30.00 | 50.00 | 66.30 | 78.00 | 46.75 | 55.00 |
| 1976 | 453,194 | 50,355 | 30.00 | 45.00 | 30.00 | 50.00 | 66.30 | 78.00 | 46.75 | 55.00 |
| 1977 | 398,323 | 44,258 | 30.00 | 45.00 | 30.00 | 50.00 | 66.30 | 78.00 | 46.75 | 55.00 |
| 1978 | 357,097 | 39,677 | 30.00 | 45.00 | 30.00 | 50.00 | 66.30 | 78.00 | 46.75 | 55.00 |
| 1979 | 318,484 | 35,387 | 30.00 | 45.00 | 30.00 | 50.00 | 66.30 | 78.00 | 46.75 | 55.00 |
| 1980 | 248,333 | 39,608 | 30.00 | 45.00 | 30.00 | 50.00 | 65.87 | 81.67 | 46.75 | 55.00 |
| 1981 | 173,667 | 32,159 | 30.00 | 45.00 | 30.00 | 50.00 | 67.55 | 83.64 | 54.25 | 63.82 |
| 1982 | 310,000 | 12,353 | 30.00 | 45.00 | 30.00 | 50.00 | 66.81 | 8277 | 51.53 | 60.62 |
| 1983 | 502,000 | 29,411 | 30.00 | 45.00 | 30.00 | 50.00 | 67.54 | 79.46 | 46.44 | 54.63 |
| 1984 | 242,666 | 19,804 | 30.00 | 45.00 | 30.00 | 50.00 | 68.68 | 80.80 | 32.19 | 37.87 |
| 1985 | 498,333 | 19,608 | 30.00 | 45.00 | 30.00 | 50.00 | 69.95 | 8229 | 34.71 | 40.83 |
| 1986 | 498,125 | 28,335 | 30.00 | 45.00 | 30.00 | 50.00 | 72.87 | 85.73 | 34.66 | 40.78 |
| 1987 | 358,842 | 27,609 | 20.00 | 40.00 | 20.00 | 45.00 | 70.66 | 83.13 | 36.27 | 42.67 |
| 1988 | 559,297 | 30,599 | 20.00 | 40.00 | 20.00 | 45.00 | 71.56 | 84.19 | 36.76 | 43.24 |
| 1989 | 305,667 | 24,891 | 20.00 | 40.00 | 20.00 | 45.00 | 68.28 | 80.33 | 45.06 | 53.01 |
| 1990 | 203,955 | 16,608 | 20.00 | 40.00 | 20.00 | 45.00 | 61.21 | 7202 | 41.15 | 48.42 |
| 1991 | 140,796 | 11,465 | 20.00 | 40.00 | 20.00 | 45.00 | 60.42 | 71.08 | 40.11 | 47.19 |
| 1992 | 219,942 | 17,910 | 20.00 | 40.00 | 20.00 | 45.00 | 58.82 | 69.20 | 38.34 | 45.11 |
| 1993 | 187,742 | 15,288 | 15.00 | 35.00 | 20.00 | 45.00 | 65.87 | 77.50 | 48.79 | 57.40 |
| 1994 | 267,928 | 21,818 | 15.00 | 35.00 | 20.00 | 45.00 | 66.73 | 78.50 | 38.59 | 45.40 |
| 1995 | 271,497 | 22,108 | 15.00 | 35.00 | 20.00 | 45.00 | 68.08 | 80.10 | 41.60 | 42.83 |
| 1996 | 230,826 | 18,797 | 15.00 | 35.00 | 15.00 | 30.00 | 66.33 | 78.03 | 37.95 | 56.18435 |
| 1997 | 194,187 | 15,813 | 10.00 | 20.00 | 10.00 | 20.00 | 64.86 | 76.30 | 46.67 | 60.00 |
| 1998 | 219,767 | 17,896 | 10.00 | 20.00 | 10.00 | 20.00 | 64.12 | 75.43 | 40.00 | 60.00 |
| 1999 | 166,887 | 13,590 | 10.00 | 20.00 | 10.00 | 20.00 | 63.50 | 74.70 | 40.00 | 62.17 |
| 2000 | 211,035 | 17,185 | 10.00 | 20.00 | 10.00 | 20.00 | 63.00 | 75.00 | 40.00 | 62.00 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| $\begin{aligned} M(\min ) & = \\ M(\max ) & = \end{aligned}$ | $\begin{aligned} & 0.005 \\ & 0.015 \end{aligned}$ |  | Retu | time (m) $=$ | 1SW(min) 1SW(max) | $\begin{aligned} & 7 \\ & 9 \end{aligned}$ | MSW(min) MSW(max) | $\begin{aligned} & 16 \\ & 18 \end{aligned}$ |  |  |

Appendix $7 f \quad$ Input data for NEAC Area Pre Fishery Abundance analysis using
Monte Carlo simulation - NORWAY-Total pre-1983

| Year | $\begin{aligned} & \text { Catch } \\ & \text { (numbers) } \end{aligned}$ |  | Unrep. as \% of total 1SW |  | Unrep. as \% of total MSW |  | $\begin{aligned} & \text { Exp. rate } \\ & \text { 1SW (\%) } \end{aligned}$ |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | min | max | min | max | min | max | min | max |
| 1971 | 213,595 | 135,247 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1972 | 279,249 | 176,818 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1973 | 305.439 | 193.402 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1974 | 288,982 | 182,981 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1975 | 271,993 | 172,224 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1976 | 270,754 | 171,439 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1977 | 263,322 | 166,733 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1978 | 285.812 | 117,655 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1979 | 324,020 | 205,167 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1980 | 323.843 | 205,055 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1981 | 221,566 | 213,943 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1982 | 163,120 | 174,229 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | 0.005 | Return time ( m ) $=$ | 1SW(min) | 7 | SW(min) | 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathrm{M}($ max $)=$ | 0.015 |  | 1SW(max) | 9 | MSW(max) | 18 |

Appendix 7g Input data for NEAC Area Pre Fishery Abundance analysis using Monte Carlo simulation - NORWAY-S (1983 onwards)

| Year | Catch (numbers) |  | Unrep. as \% of total 1SW |  | Unrep. as \% of total MSW |  | $\begin{array}{\|l\|l} \text { Exp. rate } \\ \text { 1SW (\%) } \end{array}$ |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | min | max | min | max | min | max | min | max |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 38,353 | 34,080 | 40 | 60 | 40 | 60 | 65 | 85 | 65 | 85 |
| 1984 | 86,583 | 66,580 | 40 | 60 | 40 | 60 | 65 | 85 | 65 | 85 |
| 1985 | 106,376 | 55,914 | 40 | 60 | 40 | 60 | 65 | 85 | 65 | 85 |
| 1986 | 98,525 | 76,517 | 40 | 60 | 40 | 60 | 65 | 85 | 65 | 85 |
| 1987 | 73,885 | 61,257 | 40 | 60 | 40 | 60 | 65 | 85 | 65 | 85 |
| 1988 | 72,751 | 45,278 | 40 | 60 | 40 | 60 | 65 | 85 | 65 | 85 |
| 1989 | 83,316 | 29,734 | 40 | 60 | 40 | 60 | 55 | 75 | 55 | 75 |
| 1990 | 66,477 | 38,353 | 40 | 60 | 40 | 60 | 55 | 75 | 55 | 75 |
| 1991 | 57.410 | 28,440 | 40 | 60 | 40 | 60 | 55 | 75 | 55 | 75 |
| 1992 | 40,380 | 30,691 | 40 | 60 | 40 | 60 | 55 | 75 | 55 | 75 |
| 1993 | 46,975 | 26,425 | 30 | 50 | 30 | 50 | 55 | 75 | 55 | 75 |
| 1994 | 95,265 | 31,392 | 30 | 50 | 30 | 50 | 55 | 75 | 55 | 75 |
| 1995 | 45,194 | 40,019 | 30 | 50 | 30 | 50 | 55 | 75 | 55 | 75 |
| 1996 | 22,110 | 29,417 | 30 | 50 | 30 | 50 | 55 | 75 | 55 | 75 |
| 1997 | 36.549 | 18,312 | 25 | 45 | 25 | 45 | 50 | 70 | 50 | 70 |
| 1998 | 47,909 | 23,533 | 25 | 45 | 25 | 45 | 50 | 70 | 50 | 70 |
| 1999 | 48,504 | 26,435 | 25 | 45 | 25 | 45 | 50 | 70 | 50 | 70 |
| 2000 | 85179 | 38204 | 25 | 45 | 25 | 45 | 50 | 70 | 50 | 70 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| $M($ min $)$ | 0.005 |  | Return time ( m ) $=$ |  | 1SW(min) <br> 1SW(max) | 7 | MSW(min) | 16 |  |  |
| $\mathrm{M}($ max $)$ | 0.015 |  |  |  | 9 | MSW(max) | 18 |  |  |

Appendix 7h Input data for NEAC Area Pre Fishery Abundance analysis using
Monte Carlo simulation - NORWAY-M (1983 onwards)

| Year | Catch (numbers) |  | Unrep. as \% of total 1SW |  | Unrep. as \% of total MSW |  | $\begin{aligned} & \text { Exp. rate } \\ & 1 S W(\%) \end{aligned}$ |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | min | max | min | max | min | max | min | max |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 112,811 | 73,452 | 40 | 60 | 40 | 60 | 65 | 85 | 65 | 85 |
| 1984 | 86,583 | 66,580 | 40 | 60 | 40 | 60 | 65 | 85 | 65 | 85 |
| 1985 | 106,376 | 55,914 | 40 | 60 | 40 | 60 | 65 | 85 | 65 | 85 |
| 1986 | 98,525 | 76,517 | 40 | 60 | 40 | 60 | 65 | 85 | 65 | 85 |
| 1987 | 73,885 | 61, 257 | 40 | 60 | 40 | 60 | 65 | 85 | 65 | 85 |
| 1988 | 72,751 | 45,278 | 40 | 60 | 40 | 60 | 65 | 85 | 65 | 85 |
| 1989 | 83,316 | 29,734 | 40 | 60 | 40 | 60 | 55 | 75 | 55 | 75 |
| 1990 | 66,477 | 38,353 | 40 | 60 | 40 | 60 | 55 | 75 | 55 | 75 |
| 1991 | 57,410 | 28,440 | 40 | 60 | 40 | 60 | 55 | 75 | 55 | 75 |
| 1992 | 40,380 | 30,691 | 40 | 60 | 40 | 60 | 55 | 75 | 55 | 75 |
| 1993 | 46,975 | 26,425 | 30 | 50 | 30 | 50 | 55 | 75 | 55 | 75 |
| 1994 | 95,265 | 31,392 | 30 | 50 | 30 | 50 | 55 | 75 | 55 | 75 |
| 1995 | 45,194 | 40,019 | 30 | 50 | 30 | 50 | 55 | 75 | 55 | 75 |
| 1996 | 22.110 | 29,417 | 30 | 50 | 30 | 50 | 55 | 75 | 55 | 75 |
| 1997 | 36,549 | 18,312 | 25 | 45 | 25 | 45 | 50 | 70 | 50 | 70 |
| 1998 | 47,909 | 23,533 | 25 | 45 | 25 | 45 | 50 | 70 | 50 | 70 |
| 1999 | 48,504 | 26,435 | 25 | 45 | 25 | 45 | 50 | 70 | 50 | 70 |
| 2000 | 85,179 | 38,204 | 25 | 45 | 25 | 45 | 50 | 70 | 50 | 70 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

$M($ min $)=0.005$
$M(\max )=\quad 0.015$
Return time ( $m$ )
1SW(min)
$\begin{array}{cccc}\text { 1SW(min) } & 7 & \text { MSW(min) } & 16 \\ \text { 1SW(max) } & 9 & \text { MSW(max) } & 18\end{array}$

Appendix 7i Input data for NEAC Area Pre Fishery Abundance analysis using Monte Carlo simulation - NORWAY-N (1983 onwards)

| Year | Catch (numbers) <br> 1SW |  | Unrep. as \% of total 1SW | max | Unrep. as \% of total MSW | max | $\left\lvert\, \begin{gathered}\text { Exp. rate } \\ 1 S W(\%) \\ \min \end{gathered}\right.$ | $\max$ | Exp. rate MSW (\% min | max |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1972 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1973 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1974 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 107.749 | 56,994 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1984 | 156,755 | 66,460 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1985 | 125.469 | 65,800 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1986 | 91,403 | 68,494 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1987 | 82.236 | 40,394 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1988 | 58.859 | 35.986 | 40 | 60 | 40 | 60 | 70 | 90 | 70 | 90 |
| 1989 | 53,009 | 20,466 | 40 | 60 | 40 | 60 | 60 | 80 | 60 | 80 |
| 1990 | 47,133 | 23,221 | 40 | 60 | 40 | 60 | 60 | 80 | 60 | 80 |
| 1991 | 52.442 | 30,245 | 40 | 60 | 40 | 60 | 60 | 80 | 60 | 80 |
| 1992 | 51,137 | 27,082 | 40 | 60 | 40 | 60 | 60 | 80 | 60 | 80 |
| 1993 | 56,382 | 45,669 | 30 | 50 | 30 | 50 | 60 | 80 | 60 | 80 |
| 1994 | 40.402 | 40,336 | 30 | 50 | 30 | 50 | 60 | 80 | 60 | 80 |
| 1995 | 36,706 | 30,337 | 30 | 50 | 30 | 50 | 60 | 80 | 60 | 80 |
| 1996 | 45.946 | 38.520 | 30 | 50 | 30 | 50 | 60 | 80 | 60 | 80 |
| 1997 | 43.413 | 32,701 | 25 | 45 | 25 | 45 | 60 | 80 | 60 | 80 |
| 1998 | 53,099 | 37,217 | 25 | 45 | 25 | 45 | 60 | 80 | 60 | 80 |
| 1999 | 64.531 | 49,320 | 25 | 45 | 25 | 45 | 60 | 80 | 60 | 80 |
| 2000 | 77,121 | 57.286 | 25 | 45 | 25 | 45 | 60 | 80 | 60 | 80 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

$M($ min $)=0.005$
$\begin{array}{lllll}\text { Return time }(\mathrm{m})= & \begin{array}{c}1 \mathrm{SW}(\min ) \\ 1 S W(\max )\end{array} & 7 & \mathrm{MSW}(\text { min }) & 16 \\ & \mathrm{MSW}(\text { max }) & 18\end{array}$

Appendix 7 Input data for NEAC Area Pre Fishery Abundance analysis using Monte Carlo simulation - RUSSIA.

| Year | Catch (numbers) |  | Unrep. as \% of total 1SW |  | Unrep. as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | min | max | min | max | min | max | min | max |
| 1971 | 48,312 | 80,841 | 15 | 25 | 15 | 25 | 40 | 60 | 40 | 60 |
| 1972 | 53,525 | 67,407 | 15 | 25 | 15 | 25 | 40 | 60 | 40 | 60 |
| 1973 | 89,440 | 112,636 | 15 | 25 | 15 | 25 | 40 | 60 | 40 | 60 |
| 1974 | 82,141 | 103,444 | 15 | 25 | 15 | 25 | 40 | 60 | 40 | 60 |
| 1975 | 87,944 | 129,896 | 15 | 25 | 15 | 25 | 40 | 60 | 40 | 60 |
| 1976 | 66,447 | 110,756 | 15 | 25 | 15 | 25 | 40 | 60 | 40 | 60 |
| 1977 | 55,463 | 83,195 | 15 | 25 | 15 | 25 | 40 | 60 | 40 | 60 |
| 1978 | 60,737 | 57,564 | 15 | 25 | 15 | 25 | 40 | 60 | 40 | 60 |
| 1979 | 69,423 | 63,844 | 15 | 25 | 15 | 25 | 40 | 60 | 40 | 60 |
| 1980 | 45,673 | 96,795 | 15 | 25 | 15 | 25 | 40 | 60 | 40 | 60 |
| 1981 | 32,611 | 52,528 | 15 | 25 | 15 | 25 | 40 | 60 | 40 | 60 |
| 1982 | 39,702 | 42,471 | 15 | 25 | 15 | 25 | 30 | 50 | 30 | 50 |
| 1983 | 57,870 | 68,396 | 15 | 25 | 15 | 25 | 30 | 50 | 30 | 50 |
| 1984 | 54,991 | 72,228 | 15 | 25 | 15 | 25 | 30 | 50 | 30 | 50 |
| 1985 | 72,803 | 80,292 | 15 | 25 | 15 | 25 | 30 | 50 | 30 | 50 |
| 1986 | 63,926 | 89,465 | 15 | 25 | 15 | 25 | 30 | 50 | 30 | 50 |
| 1987 | 97,242 | 41,769 | 15 | 25 | 15 | 25 | 30 | 45 | 30 | 50 |
| 1988 | 53,158 | 46,848 | 15 | 25 | 15 | 25 | 30 | 45 | 30 | 50 |
| 1989 | 78,023 | 29,454 | 15 | 25 | 20 | 30 | 30 | 45 | 15 | 30 |
| 1990 | 70,595 | 25,663 | 15 | 25 | 20 | 30 | 30 | 45 | 15 | 25 |
| 1991 | 40,603 | 17,543 | 33 | 47 | 33 | 47 | 25 | 35 | 15 | 25 |
| 1992 | 34,021 | 13,431 | 35 | 45 | 45 | 55 | 25 | 35 | 15 | 25 |
| 1993 | 28,100 | 17,907 | 35 | 45 | 50 | 60 | 25 | 35 | 15 | 25 |
| 1994 | 30,877 | 13,668 | 35 | 45 | 55 | 65 | 25 | 35 | 15 | 25 |
| 1995 | 27,775 | 10,023 | 35 | 45 | 55 | 65 | 25 | 35 | 15 | 25 |
| 1996 | 33,878 | 8,708 | 35 | 45 | 65 | 75 | 25 | 35 | 10 | 20 |
| 1997 | 31,857 | 7,107 | 35 | 45 | 65 | 75 | 25 | 35 | 10 | 20 |
| 1998 | 34,870 | 7,024 | 35 | 45 | 65 | 75 | 25 | 35 | 10 | 20 |
| 1999 | 24,016 | 6,998 | 35 | 45 | 65 | 75 | 25 | 35 | 10 | 20 |
| 2000 | 27,702 | 9,413 | 35 | 45 | 65 | 75 | 25 | 35 | 10 | 20 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

$M($ min $)=0.005$
$M(\max )=0.015$

Return time $(\mathrm{m})=$
1SW(min)
1SW(max)

MSW(min)
16

## Appendix 7k Input data for NEAC Area Pre Fishery Abundance analysis using Mbnte Carlo simulation-SWEDEN

| Year | Catch(numbers) |  | Unrep. as \%of total 1SW |  | Unrep. as \%of total MSW |  | $\begin{array}{\|l\|l} \text { Exp. rate } \\ \text { 1SW(\%) } \end{array}$ |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | min | max | min | max | min | max | min | max |
| 1971 | 6,330 | 420 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1972 | 5,005 | 295 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1973 | 6,210 | 1,025 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1974 | 8,935 | 660 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1975 | 9,620 | 160 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1976 | 5,420 | 480 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1977 | 2,555 | 360 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1978 | 2917 | 275 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1979 | 3,080 | 800 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1980 | 3,920 | 1,400 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1981 | 7,095 | 407 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1982 | 6,230 | 1,460 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1983 | 8,290 | 1,005 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1984 | 11,680 | 1,410 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1985 | 13,890 | 590 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1986 | 14,635 | 570 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1987 | 11,860 | 1,700 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1988 | 9,930 | 1,650 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1989 | 3,180 | 4,610 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1990 | 7,430 | 3,135 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1991 | 8,990 | 3,620 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1992 | 9,850 | 4,655 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1993 | 10,540 | 6,370 | 20 | 50 | 20 | 50 | 70 | 95 | 55 | 100 |
| 1994 | 8,035 | 4,660 | 20 | 50 | 20 | 50 | 60 | 85 | 55 | 100 |
| 1995 | 9,761 | 2.770 | 20 | 50 | 20 | 50 | 50 | 75 | 55 | 90 |
| 1996 | 6,008 | 3,542 | 20 | 50 | 20 | 50 | 50 | 75 | 55 | 90 |
| 1997 | 2747 | 2,307 | 20 | 50 | 20 | 50 | 50 | 75 | 55 | 90 |
| 1998 | 2421 | 1,702 | 5 | 25 | 5 | 25 | 60 | 85 | 55 | 90 |
| 1999 | 3,573 | 1,460 | 5 | 25 | 5 | 25 | 55 | 90 | 55 | 90 |
| 2000 | 7,103 | 3,196 | 5 | 25 | 5 | 25 | 55 | 90 | 55 | 90 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

$\begin{aligned} M(\min ) & =0.005 \\ M(\max ) & =0.015\end{aligned}$
Retumtime $(m)=$
1SMmin)
1SW(max
$\begin{array}{lrl}7 & \text { MSW(min) } & 16 \\ 9 & \text { MSW (max) } & 18\end{array}$

Appendix 71 Input data for NEAC Area Pre Fishery Abundance analysis using Monte Carlo simulation - UK(England and Wales).

| Year | Catch (numbers) |  | Unrep. as \% of total 1SW |  | Unrep. as \% of total MSW |  | $\begin{array}{\|l\|l} \text { Exp. rate } \\ \text { 1SW (\%) } \end{array}$ |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | min | max | min | max | min | max | min | max |
| 1971 | 34,142 | 18,384 | 25 | 45 | 25 | 45 | 23 | 43 | 18 | 38 |
| 1972 | 38,630 | 20,347 | 25 | 45 | 25 | 45 | 23 | 43 | 18 | 38 |
| 1973 | 36,357 | 18,729 | 25 | 45 | 25 | 45 | 23 | 43 | 18 | 38 |
| 1974 | 36,048 | 18,159 | 25 | 45 | 25 | 45 | 23 | 43 | 18 | 38 |
| 1975 | 41,730 | 20,554 | 25 | 45 | 25 | 45 | 23 | 43 | 18 | 38 |
| 1976 | 26,686 | 12,849 | 25 | 45 | 25 | 45 | 24 | 44 | 18 | 38 |
| 1977 | 30,665 | 14,431 | 25 | 45 | 25 | 45 | 24 | 44 | 19 | 39 |
| 1978 | 31,253 | 14,372 | 25 | 45 | 25 | 45 | 25 | 45 | 19 | 39 |
| 1979 | 24,725 | 11,108 | 25 | 45 | 25 | 45 | 26 | 46 | 20 | 40 |
| 1980 | 34,066 | 14,950 | 25 | 45 | 25 | 45 | 29 | 49 | 23 | 43 |
| 1981 | 41,670 | 17,859 | 25 | 45 | 25 | 45 | 29 | 49 | 23 | 43 |
| 1982 | 25,480 | 10,662 | 25 | 45 | 25 | 45 | 30 | 50 | 24 | 44 |
| 1983 | 35,332 | 14,432 | 25 | 45 | 25 | 45 | 31 | 51 | 24 | 44 |
| 1984 | 30,943 | 12,334 | 25 | 45 | 25 | 45 | 32 | 52 | 25 | 45 |
| 1985 | 33,680 | 13,098 | 25 | 45 | 25 | 45 | 33 | 53 | 26 | 46 |
| 1986 | 41,700 | 15,817 | 25 | 45 | 25 | 45 | 33 | 53 | 26 | 46 |
| 1987 | 36,339 | 17,101 | 25 | 45 | 25 | 45 | 33 | 53 | 26 | 46 |
| 1988 | 47,242 | 21,225 | 25 | 45 | 25 | 45 | 32 | 52 | 26 | 46 |
| 1989 | 32,559 | 17,532 | 20 | 40 | 20 | 40 | 34 | 54 | 28 | 48 |
| 1990 | 23,635 | 21,817 | 20 | 40 | 20 | 40 | 35 | 55 | 28 | 48 |
| 1991 | 22,408 | 9,152 | 20 | 40 | 20 | 40 | 34 | 54 | 28 | 48 |
| 1992 | 22,233 | 6,641 | 20 | 40 | 20 | 40 | 34 | 54 | 27 | 47 |
| 1993 | 29,963 | 7,028 | 30 | 60 | 30 | 60 | 29 | 49 | 23 | 43 |
| 1994 | 40,610 | 12,130 | 30 | 60 | 30 | 60 | 28 | 48 | 22 | 42 |
| 1995 | 29,211 | 11,360 | 15 | 25 | 15 | 25 | 27 | 47 | 21 | 41 |
| 1996 | 21,415 | 11,531 | 15 | 25 | 15 | 25 | 23 | 43 | 18 | 38 |
| 1997 | 18,521 | 6,850 | 15 | 25 | 15 | 25 | 22 | 42 | 17 | 37 |
| 1998 | 19,726 | 4,040 | 15 | 25 | 15 | 25 | 20 | 40 | 15 | 35 |
| 1999 | 15,430 | 6,613 | 15 | 25 | 15 | 25 | 17 | 37 | 12 | 32 |
| 2000 | 23,461 | 7,008 | 15 | 25 | 15 | 25 | 16 | 36 | 12 | 32 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| $\begin{array}{r} M(\min )= \\ M(\max )= \end{array}$ | 0.005 0.015 |  | Retu | ime (m | 1SW(min) 1SW(max) | 7 9 | MSW(min) MSW(max) | 17 19 |  |  |

Appendix 7m Input data for NEAC Area Pre Fishery Abundance analysis using
Monte Carlo simulation - UK(Northern Ireland)-1

| Year | Catch (numbers) |  | Unrep. as \% of total 1SW |  | Unrep. as \% of total MSW |  | Exp. rate 1SW (\%) |  | Exp. rateMSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | min | max | min | max | min | max | min | max |
| 1971 | 79.715 | 4,196 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1972 | 66,054 | 3,477 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1973 | 58.705 | 3.090 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1974 | 74,148 | 3,903 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1975 | 52,159 | 2.745 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1976 | 36,984 | 1.947 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1977 | 37,295 | 1,963 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1978 | 45,515 | 2,396 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1979 | 35,153 | 1,850 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1980 | 46,762 | 2.461 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1981 | 33.042 | 1,739 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1982 | 57,149 | 3.008 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1983 | 79,089 | 4,163 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1984 | 28.055 | 1.477 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1985 | 38.495 | 2.026 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1986 | 44,036 | 2.318 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1987 | 17,559 | 924 | 10 | 33 | 10 | 33 | 62 | 76 | 41 | 51 |
| 1988 | 44,920 | 2,364 | 10 | 33 | 10 | 33 | 58 | 71 | 32 | 40 |
| 1989 | 61,585 | 3.241 | 10 | 37 | 10 | 37 | 80 | 98 | 54 | 66 |
| 1990 | 40.732 | 2.144 | 10 | 17 | 10 | 17 | 56 | 68 | 34 | 42 |
| 1991 | 22.176 | 1,167 | 10 | 17 | 10 | 17 | 58 | 71 | 39 | 47 |
| 1992 | 40,144 | 2.113 | 10 | 23 | 10 | 23 | 50 | 62 | 30 | 36 |
| 1993 | 36,127 | 1.901 | 10 | 17 | 10 | 17 | 37 | 45 | 11 | 13 |
| 1994 | 36,921 | 1,943 | 10 | 28 | 10 | 28 | 63 | 77 | 36 | 44 |
| 1995 | 34,116 | 1,796 | 10 | 17 | 10 | 17 | 60 | 74 | 38 | 46 |
| 1996 | 29,017 | 1.527 | 10 | 20 | 10 | 20 | 47 | 67 | 24 | 44 |
| 1997 | 41,765 | 2,198 | 5 | 15 | 5 | 15 | 50 | 70 | 24 | 44 |
| 1998 | 37,953 | 1,998 | 5 | 15 | 5 | 15 | 20 | 30 | 15 | 30 |
| 1999 | 22,126 | 1,165 | 5 | 15 | 5 | 15 | 58 | 68 | 25 | 40 |
| 2000 | 31,038 | 1,634 | 5 | 15 | 5 | 15 | 53 | 63 | 25 | 40 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

$\begin{array}{ll}M(\min ) & =0.005 \\ M(\max ) & =0.015\end{array}$

1SW(max) $\quad 9 \quad \mathrm{MSW}($ max $) \quad 18$

Appendix 7n
Input data for NEAC Area Pre Fishery Abundance analysis using
Monte Carlo simulation - UK(Northern Ireland)-FCB area-(2)

| Year | $\begin{aligned} & \text { Catch } \\ & \text { (numbers) } \end{aligned}$ |  | $\begin{aligned} & \text { Unrep. as } \\ & \% \text { of total } \\ & 1 S W \end{aligned}$ |  | $\begin{aligned} & \text { Unrep. as } \\ & \text { \%of total } \\ & \text { MSW } \end{aligned}$ |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1SW | MSW | min | max | min | max | min | max | min | max |
| 1971 | 36,270 | 1.909 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1972 | 35,293 | 1,858 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1973 | 29,858 | 1,571 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1974 | 22,787 | 1,199 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1975 | 27,275 | 1.436 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1976 | 18.270 | 962 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1977 | 17,139 | 902 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1978 | 25,391 | 1,336 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1979 | 14,631 | 70 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1980 | 16,310 | 858 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1981 | 16,338 | 860 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1982 | 14,370 | 756 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1983 | 21,293 | 1,121 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1984 | 11,348 | 597 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1985 | 12.635 | 665 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1986 | 13.443 | 708 | 10 | 33 | 10 | 33 | 75 | 85 | 45 | 55 |
| 1987 | 9,439 | 497 | 10 | 33 | 10 | 33 | 62 | 76 | 41 | 51 |
| 1988 | 14,628 | 70 | 10 | 33 | 10 | 33 | 58 | 71 | 32 | 40 |
| 1989 | 15.405 | 811 | 10 | 37 | 10 | 37 | 80 | 98 | 54 | 66 |
| 1990 | 10,215 | 538 | 10 | 17 | 10 | 17 | 56 | 68 | 34 | 42 |
| 1991 | 6.804 | 358 | 10 | 17 | 10 | 17 | 58 | 71 | 39 | 47 |
| 1992 | 9,534 | 502 | 10 | 23 | 10 | 23 | 50 | 62 | 30 | 36 |
| 1993 | 8.939 | 470 | 10 | 17 | 10 | 17 | 37 | 45 | 11 | 13 |
| 1994 | 11,146 | 587 | 10 | 28 | 10 | 28 | 63 | 77 | 36 | 44 |
| 1995 | 11,887 | 626 | 10 | 17 | 10 | 17 | 60 | 74 | 38 | 46 |
| 1996 | 10.606 | 558 | 10 | 20 | 10 | 20 | 47 | 67 | 24 | 44 |
| 1997 | 10,705 | 563 | 5 | 15 | 5 | 15 | 50 | 70 | 24 | 44 |
| 1998 | 9,577 | 504 | 5 | 15 | 5 | 15 | 20 | 30 | 15 | 30 |
| 1999 | 9,205 | 484 | 5 | 15 | 5 | 15 | 58 | 68 | 25 | 40 |
| 2000 | 10,826 | 570 | 5 | 15 | 5 | 15 | 53 | 63 | 25 | 40 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

$\begin{array}{ll}M(\text { min })= & 0.005 \\ M(\text { max })= & 0.015\end{array}$
$M(\max )=0.015$
$\begin{array}{lllll}\text { Return time }(\mathrm{m})= & \begin{array}{c}\text { 1SW(min) } \\ \text { 1SW (max) }\end{array} & 7 & \text { MSW(min) } & 16 \\ \text { MSW(max) } & 18\end{array}$

Appendx 70 Input data for NEAC Area Pre Fishery Abundance analysis using Mbnte Carlo simulation - UK(Scotland)-E

| Year | Catch (numbers) 1SW | MSW | Catch of Sootish fish in Enaland (\%1SM | Unrep. as <br> \%of total <br> 1SW | max | Unrep. as \%of total MSW min | max | $\begin{gathered} \text { Exp. rate } \\ \text { 1SW(\%) } \\ \text { min } \end{gathered}$ | max | Exp. rate MSW(\%) <br> min | max |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 70\% |  |  |  |  |  |  |  |  |
| 1971 | 216,873 | 135.527 | 5,335 | 15 | 35 | 15 | 35 | 30.0 | 50.0 | 40.0 | 60.0 |
| 1972 | 220,106 | 183,872 | 49,097 | 15 | 35 | 15 | 35 | 29.3 | 49.1 | 39.0 | 58.7 |
| 1973 | 259,773 | 204,825 | 59,700 | 15 | 35 | 15 | 35 | 28.7 | 48.1 | 38.0 | 57.4 |
| 1974 | 245,424 | 158,951 | 50,118 | 15 | 35 | 15 | 35 | 28.0 | 47.2 | 37.0 | 56.2 |
| 1975 | 181,940 | 180,828 | 50,778 | 15 | 35 | 15 | 35 | 27.4 | 46.3 | 36.0 | 54.9 |
| 1976 | 150,069 | 92.179 | 14,759 | 15 | 35 | 15 | 35 | 26.7 | 45.3 | 35.0 | 53.6 |
| 197 | 154,306 | 118.645 | 49,186 | 15 | 35 | 15 | 35 | 26.1 | 44.4 | 34.0 | 52.3 |
| 1978 | 158,844 | 139,688 | 47,500 | 15 | 35 | 15 | 35 | 25.4 | 43.5 | 33.0 | 51.1 |
| 1979 | 160,791 | 116,514 | 39,552 | 15 | 35 | 15 | 35 | 24.8 | 42.6 | 32.0 | 49.8 |
| 1980 | 101,665 | 155.646 | 41,202 | 10 | 25 | 10 | 25 | 24.1 | 41.6 | 31.0 | 48.5 |
| 1981 | 129,690 | 156.683 | 61,511 | 10 | 25 | 10 | 25 | 23.4 | 40.7 | 30.0 | 47.2 |
| 1982 | 175,355 | 113,180 | 44,147 | 10 | 25 | 10 | 25 | 228 | 39.8 | 29.0 | 46.0 |
| 1983 | 170,843 | 126.104 | 67,231 | 10 | 25 | 10 | 25 | 221 | 38.8 | 28.0 | 44.7 |
| 1984 | 175,675 | 90,829 | 50,994 | 10 | 25 | 10 | 25 | 21.5 | 37.9 | 27.0 | 43.4 |
| 1985 | 133,073 | 95,012 | 48,753 | 10 | 25 | 10 | 25 | 20.8 | 37.0 | 26.0 | 42.1 |
| 1986 | 180,276 | 128,813 | 53,277 | 10 | 25 | 10 | 25 | 20.2 | 36.0 | 25.0 | 40.9 |
| 1987 | 139.252 | 88.519 | 20,999 | 10 | 25 | 10 | 25 | 19.5 | 35.1 | 24.0 | 39.6 |
| 1988 | 118,580 | 91,068 | 41,696 | 10 | 25 | 10 | 25 | 18.9 | 34.2 | 23.0 | 38.3 |
| 1989 | 142,992 | 85,348 | 33,577 | 5 | 15 | 5 | 15 | 18.2 | 33.2 | 22.0 | 37.0 |
| 1990 | 63,297 | 73,954 | 41,224 | 5 | 15 | 5 | 15 | 17.6 | 32.3 | 21.0 | 35.8 |
| 1991 | 53,835 | 53,676 | 20,343 | 5 | 15 | 5 | 15 | 16.9 | 31.4 | 20.0 | 34.5 |
| 1992 | 79,883 | 67,968 | 16.115 | 5 | 15 | 5 | 15 | 16.2 | 30.4 | 19.0 | 33.2 |
| 1993 | 73,396 | 60,496 | 33,440 | 5 | 15 | 5 | 15 | 15.6 | 29.5 | 18.0 | 31.9 |
| 1994 | 80,555 | 72746 | 37,243 | 5 | 15 | 5 | 15 | 14.9 | 28.6 | 17.0 | 30.7 |
| 1995 | 72986 | 69.115 | 42,568 | 5 | 15 | 5 | 15 | 14.3 | 27.7 | 16.0 | 29.4 |
| 1996 | 56,617 | 50,361 | 14,865 | 5 | 15 | 5 | 15 | 13.6 | 26.7 | 15.0 | 28.1 |
| 1997 | 37,465 | 34,841 | 17,538 | 5 | 15 | 5 | 15 | 13.0 | 25.8 | 14.0 | 26.8 |
| 1998 | 44,915 | 32,264 | 14,612 | 5 | 15 | 5 | 15 | 123 | 24.9 | 13.0 | 25.6 |
| 1999 | 20.840 | 26,979 | 21,466 | 5 | 15 | 5 | 15 | 11.7 | 23.9 | 12.0 | 24.3 |
| 2000 | 21,328 | 25.287 | 34,666 | 5 | 15 | 5 | 15 | 11.0 | 23.0 | 11.0 | 23.0 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

$M($ min $)=0.005$
$M(\max )=0.015$

Appendix 7p Input data for NEAC Area Pre Fishery Abundance analysis using
Monte Carlo simulation - UK(Scotland)-W.

| Year | Catch (numbers) |  | Unrep. as <br> \% of total <br> 1SW |  | Unrep. as <br> \% of total <br> MSW |  | Exp. rate 1SW (\%) |  | Exp. rate MSW (\%) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1sw | MSW | min | max | min | max | min | max | min | max |
| 1971 | 45287 | 26074 | 25 | 45 | 25 | 45 | 10 | 30 | 20 | 40 |
| 1972 | 31359 | 34151 | 25 | 45 | 25 | 45 | 10 | 30 | 19 | 39 |
| 1973 | 33317 | 33095 | 25 | 45 | 25 | 45 | 10 | 29 | 19 | 38 |
| 1974 | 43992 | 29406 | 25 | 45 | 25 | 45 | 9 | 29 | 18 | 38 |
| 1975 | 40424 | 27150 | 25 | 45 | 25 | 45 | 9 | 28 | 18 | 37 |
| 1976 | 38423 | 22403 | 25 | 45 | 25 | 45 | 9 | 28 | 17 | 36 |
| 1977 | 39958 | 20342 | 25 | 45 | 25 | 45 | 9 | 27 | 17 | 35 |
| 1978 | 45626 | 23266 | 25 | 45 | 25 | 45 | 9 | 27 | 16 | 34 |
| 1979 | 26445 | 15995 | 25 | 45 | 25 | 45 | 9 | 26 | 16 | 34 |
| 1980 | 19776 | 16942 | 20 | 35 | 20 | 35 | 8 | 26 | 15 | 33 |
| 1981 | 21048 | 18038 | 20 | 35 | 20 | 35 | 8 | 26 | 15 | 32 |
| 1982 | 32706 | 15062 | 20 | 35 | 20 | 35 | 8 | 25 | 14 | 31 |
| 1983 | 38774 | 19857 | 20 | 35 | 20 | 35 | 8 | 25 | 14 | 30 |
| 1984 | 37404 | 16384 | 20 | 35 | 20 | 35 | 8 | 24 | 13 | 30 |
| 1985 | 24939 | 19636 | 20 | 35 | 20 | 35 | 8 | 24 | 13 | 29 |
| 1986 | 22579 | 19584 | 20 | 35 | 20 | 35 | 7 | 23 | 12 | 28 |
| 1987 | 25533 | 15475 | 20 | 35 | 20 | 35 | 7 | 23 | 12 | 27 |
| 1988 | 30518 | 21094 | 20 | 35 | 20 | 35 | 7 | 22 | 11 | 27 |
| 1989 | 31949 | 18538 | 15 | 25 | 15 | 25 | 7 | 22 | 11 | 26 |
| 1990 | 17797 | 13970 | 15 | 25 | 15 | 25 | 7 | 21 | 10 | 25 |
| 1991 | 19773 | 11517 | 15 | 25 | 15 | 25 | 7 | 21 | 10 | 24 |
| 1992 | 21793 | 14873 | 15 | 25 | 15 | 25 | 6 | 21 | 9 | 23 |
| 1993 | 21121 | 11230 | 15 | 25 | 15 | 25 | 6 | 20 | 9 | 23 |
| 1994 | 18904 | 12658 | 15 | 25 | 15 | 25 | 6 | 20 | 8 | 22 |
| 1995 | 16935 | 9337 | 15 | 25 | 15 | 25 | 6 | 19 | 8 | 21 |
| 1996 | 9796 | 7559 | 15 | 25 | 15 | 25 | 6 | 19 | 7 | 20 |
| 1997 | 9407 | 5586 | 15 | 25 | 15 | 25 | 6 | 18 | 7 | 19 |
| 1998 | 8532 | 6984 | 15 | 25 | 15 | 25 | 5 | 18 | 6 | 19 |
| 1999 | 4343 | 3672 | 15 | 25 | 15 | 25 | 5 | 17 | 6 | 18 |
| 2000 | 5568 | 4487 | 15 | 25 | 15 | 25 | 5 | 17 | 5 | 17 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| $M($ min $)=$ <br> $M(\max )=$ | $\begin{aligned} & 0.005 \\ & 0.015 \end{aligned}$ |  | Retu | ime ( $m$ ) $=$ | $\begin{aligned} & \text { 1SW(min) } \\ & \text { 1SWW(max) } \end{aligned}$ | 7 9 | MSW(min) MSW(max) | 16 18 |  |  |

Appendix 7q Input data for NEAC Area Pre Fishery Abundance analysis using Monte Carlo simulation - FAROES

| $\left\lvert\, \begin{aligned} & \text { Year } \\ & n / n+1 \end{aligned}\right.$ | Catch (numbers) 1SW | MSW | Unrep. as \% of total 1SW <br> min | max | Unrep. as \% of total MSW <br> min | max | Exp. rate 1SW (\%) <br> min | max | Exp. rate MSW (\%) <br> min | max |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 2620 | 105796 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1972 | 2754 | 111187 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1973 | 3121 | 126012 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1974 | 2186 | 88276 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1975 | 2798 | 112984 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1976 | 1830 | 73900 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1977 | 1291 | 52112 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1978 | 974 | 39309 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1979 | 1736 | 70082 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1980 | 4523 | 182616 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1981 | 7443 | 300542 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1982 | 6859 | 276957 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1983 | 15861 | 215349 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1984 | 5534 | 138227 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1985 | 378 | 158103 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1986 | 1979 | 180934 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1987 | 90 | 166244 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1988 | 8637 | 87629 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1989 | 1788 | 121965 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1990 | 1989 | 140054 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1991 | 943 | 84935 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1992 | 68 | 35700 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1993 | 6 | 30023 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1994 | 15 | 31672 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1995 | 18 | 34662 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1996 | 101 | 28381 | 5 | 15 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1997 | 0 | 0 | 10 | 20 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1998 | 339 | 1,424 | 10 | 20 | 0 | 0 | 100 | 100 | 100 | 100 |
| 1999 | 0 | 0 | 10 | 20 | 0 | 0 | 100 | 100 | 100 | 100 |
| 2000 | 225 | 1,765 | 10 | 20 | 0 | 0 | 100 | 100 | 100 | 100 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 100 | 100 | 100 | 100 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 100 | 100 | 100 | 100 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 100 | 100 | 100 | 100 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 100 | 100 | 100 | 100 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 100 | 100 | 100 | 100 |

$M(\min )=0.005$

Return time $(\mathrm{m})=$| $1 \mathrm{SW}($ min $)$ | 0 | $\mathrm{MSW}($ min $)$ | 1 |
| :---: | :---: | :---: | :---: |
| $1 \mathrm{SW}(\max )$ | 1 | $\mathrm{MSW}(\max )$ | 2 |

```
min 0.170
    max 0.270
```

Appendix 7r Input data for NEAC Area Pre Fishery Abundanoe analysis using Mbnte Carlo simulation- WEST GR $\rightarrow$ NAND.

| Year | $\begin{gathered} \begin{array}{c} \text { Catch } \\ \text { (numbers) } \end{array} \\ \\ 1 \mathrm{SW} \\ \hline \end{gathered}$ | MSW | Uurep. as <br> \%of total <br> 1SW <br> min |  | Unrep. as \%of total MSW <br> min | max | Exp. rate 1SW(\%) <br> min | max | Exp. rate MSW(\%) <br> min | max |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 0 | 856369 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1972 | 0 | 614244 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1973 | 0 | 560048 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1974 | 0 | 535475 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1975 | 0 | 650641 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1976 | 0 | 386513 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1977 | 0 | 442368 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1978 | 0 | 293731 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1979 | 0 | 417665 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1980 | 0 | 370807 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1981 | 0 | 398738 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1982 | 0 | 346302 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1983 | 0 | 100000 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1984 | 0 | 95498 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1985 | 0 | 301045 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1986 | 0 | 316832 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1987 | 0 | 305696 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1988 | 0 | 280818 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1989 | 0 | 117422 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1990 | 0 | 101859 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1991 | 0 | 178113 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1992 | 0 | 84342 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1993 | 0 | 2,000 | 0 | 0 | -25 | 25 | 100 | 100 | 100 | 100 |
| 1994 | 0 | 2,000 | 0 | 0 | -25 | 25 | 100 | 100 | 100 | 100 |
| 1995 | 0 | 32422 | 0 | 0 | 5 | 15 | 100 | 100 | 100 | 100 |
| 1996 | 0 | 31944 | 0 | 0 | 10 | 20 | 100 | 100 | 100 | 100 |
| 1997 | 0 | 21402 | 0 | 0 | 9 | 19 | 100 | 100 | 100 | 100 |
| 1998 | 0 | 395 | 0 | 0 | 3 | 13 | 100 | 100 | 100 | 100 |
| 1999 | 0 | 6169 | 0 | 0 | 40 | 60 | 100 | 100 | 100 | 100 |
| 2000 | 0 | 8171 | 0 | 0 | 30 | 50 | 100 | 100 | 100 | 100 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 100 | 100 | 100 | 100 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 100 | 100 | 100 | 100 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 100 | 100 | 100 | 100 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 100 | 100 | 100 | 100 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 100 | 100 | 100 | 100 |
| $\begin{aligned} M(\min ) & = \\ M(\max ) & = \end{aligned}$ | $\begin{aligned} & 0.005 \\ & 0.015 \end{aligned}$ |  | Retu | me(m) | $\begin{gathered} \text { 1SW(min) } \\ \text { 1SW(max) } \end{gathered}$ | 7 | MSW(min) <br> MSW(max) | 8 10 |  |  |

Appendix 8a Lagged egg deposition analysis and estimation of conservation limit options - FINLAND

|  | Est. 1SW spawners | Est MSW spawners | Egg deposition $\operatorname{egg} \times 10^{-3}$ | Smolt age composition |  |  |  |  |  | $\begin{gathered} \text { Lagged } \\ \text { egg dep. } \\ \mathrm{S} \\ \mathrm{egg} \times 10^{-3} \\ \hline \end{gathered}$ | Total 1SW recruits R | R/S |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Egg | 5000 | 13000 |  | 1 yr | 2 yr | 3 yr | 4 yr | 5 yr | 6 yr |  |  |  |
| Fem | 12\% | 77\% |  | 0.00 | 0.00 | 0.26 | 0.59 | 0.14 | 0.01 |  |  |  |
| 1971 | 3,618 | 3,083 | 33,027 |  |  |  |  |  |  | n/a | 22,643 |  |
| 1972 | 5.557 | 4.835 | 51,737 |  |  |  |  |  |  | n/a | 38,197 |  |
| 1973 | 8,399 | 7,422 | 79,331 |  |  |  |  |  |  | n/a | 43,245 |  |
| 1974 | 7.821 | 7.100 | 75,762 | 0 |  |  |  |  |  | n/a | 41,753 |  |
| 1975 | 7.825 | 6.875 | 73.510 | 0 | 0 |  |  |  |  | n/a | 38,643 |  |
| 1976 | 6,722 | 6,149 | 65,581 | 0 | 0 | 8,587 |  |  |  | n/a | 33,478 |  |
| 1977 | 6,096 | 5,672 | 60,434 | 0 | 0 | 13,452 | 19,486 |  |  | n/a | 25,723 |  |
| 1978 | 4.429 | 3.455 | 37,242 | 0 | 0 | 20,626 | 30,525 | 4.624 |  | n/a | 20,305 |  |
| 1979 | 4.553 | 3,396 | 36,726 | 0 | 0 | 19.698 | 46,805 | 7.243 | 330 | 74.077 | 26,369 | 0.36 |
| 1980 | 4,501 | 4,617 | 48,919 | 0 | 0 | 19,113 | 44,699 | 11,106 | 517 | 75,436 | 35,161 | 0.47 |
| 1981 | 4,173 | 7,191 | 74,485 | 0 | 0 | 17,051 | 43,371 | 10,607 | 793 | 71,822 | 35,731 | 0.50 |
| 1982 | 2,979 | 7,983 | 81,700 | 0 | 0 | 15,713 | 38,693 | 10,291 | 758 | 65,455 | 33,385 | 0.51 |
| 1983 | 4,539 | 8,837 | 91,181 | 0 | 0 | 9,683 | 35,656 | 9,181 | 735 | 55,255 | 31,005 | 0.56 |
| 1984 | 5,685 | 7.177 | 75,252 | 0 | 0 | 9,549 | 21,973 | 8.461 | 656 | 40,638 | 33,843 | 0.83 |
| 1985 | 7.137 | 7.219 | 76.549 | 0 | 0 | 12,719 | 21,669 | 5.214 | 604 | 40,206 | 32,418 | 0.81 |
| 1986 | 7,175 | 4,840 | 52,756 | 0 | 0 | 19,366 | 28,862 | 5,142 | 372 | 53,742 | 36,580 | 0.68 |
| 1987 | 10,019 | 6,766 | 73,734 | 0 | 0 | 21,242 | 43,946 | 6,849 | 367 | 72,404 | 38,276 | 0.53 |
| 1988 | 6,820 | 5,391 | 58,052 | 0 | 0 | 23,707 | 48,203 | 10,428 | 489 | 82,827 | 34,390 | 0.42 |
| 1989 | 7.787 | 5,303 | 57,758 | 0 | 0 | 19,566 | 53,797 | 11.438 | 745 | 85,545 | 42.054 | 0.49 |
| 1990 | 7,804 | 5,506 | 59,798 | 0 | 0 | 19,903 | 44,399 | 12,765 | 817 | 77,884 | 41,788 | 0.54 |
| 1991 | 7,010 | 6,155 | 65,817 | 0 | 0 | 13,717 | 45,164 | 10,535 | 912 | 70,327 | 38,955 | 0.55 |
| 1992 | 11,433 | 6.092 | 67,837 | 0 | 0 | 19,171 | 31,126 | 10,717 | 753 | 61,766 | 57.028 | 0.92 |
| 1993 | 8,430 | 7,845 | 83,583 | 0 | 0 | 15,094 | 43,503 | 7,386 | 765 | 66,748 | 42,892 | 0.64 |
| 1994 | 5,699 | 6,038 | 63,855 | 0 | 0 | 15,017 | 34,251 | 10,323 | 528 | 60,118 | 32,391 | 0.54 |
| 1995 | 5,852 | 5,258 | 56,148 | 0 | 0 | 15,548 | 34,077 | 8,127 | 737 | 58,489 | 27,377 | 0.47 |
| 1996 | 12,435 | 4,065 | 48,149 | 0 | 0 | 17,112 | 35,281 | 8,086 | 581 | 61,060 | 43,504 | 0.71 |
| 1997 | 10,997 | 6.413 | 70,793 | 0 | 0 | 17.638 | 38,832 | 8.372 | 578 | 65.419 | 38.418 | 0.59 |
| 1998 | 13,201 | 5,710 | 65,078 | 0 | 0 | 21,732 | 40,024 | 9,214 | 598 | 71,568 | 42,296 | 0.59 |
| 1999 | 12,688 | 4,506 | 52,714 | 0 | 0 | 16,602 | 49,314 | 9,497 | 658 | 76,072 | 64,082 | 0.84 |
| 2000 | 13,463 | 10,202 | 110,199 | 0 | 0 | 14,598 | 37,674 | 11,702 | 678 | 64,653 | 39,935 | 0.62 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 12,519 | 33,127 | 8,940 | 836 | 55,422 | 0 | 0.00 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 18,406 | 28,408 | 7,861 | 639 | 55,314 | 0 | 0.00 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 16,920 | 41,768 | 6.741 | 561 | 65,991 | 0 | 0.00 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 13,706 | 38,396 | 9,911 | 481 | 62,494 | 0 | 0.00 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 28.652 | 31,101 | 9,111 | 708 | 69,572 | 0 | 0.00 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 65,017 | 7,380 | 651 | 73,048 | 0 | 0.00 |




| Conservation limit: |  |  |  |
| :---: | :---: | :---: | :---: |
| Eggs /1000 | 1SW | MSW | Total salmon |
| 61,593 | 9,112 | 5,607 | 14,718 |

Appendix 8b Lagged egg deposition analysis and estimation of conservation limit options - ICELAND-1

|  | Est. 1SW spawners | Est MSW spawners | $\left\|\begin{array}{c} \text { Egg } \\ \text { deposition } \\ \text { egg } \times 10^{-3} \end{array}\right\|$ | Smolt age composition |  |  |  |  |  | $\begin{array}{\|c\|} \hline \text { Lagged } \\ \text { egg dep. } \\ S \\ \mathrm{~S} \\ \hline \mathrm{egg} \times 10^{-3} \\ \hline \end{array}$ | Total 1SW recruits R | R/S | Quasi S-R Relationship <br> 120,000 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Egg | 5725 | 10256 |  | 1 yr | 2 yr | 3 yr | 4 yr | 5 yr | 6 yr |  |  |  | 100 |  | - |  |  |
| Fem | 56\% | 62\% |  | 0.08 | 0.49 | 0.35 | 0.08 | 0.00 | 0.00 |  |  |  | 100,000 |  |  | , |  |
| 1971 | 27.188 | 10,673 | 155,031 |  |  |  |  |  |  | n/a | 98,732 |  | 80,000 |  | $\triangle$ |  |  |
| 1972 | 27,937 | 13,684 | 176,579 |  |  |  |  |  |  | n/a | 98,849 |  | \% 60 |  | 24 | 4 |  |
| 1973 | 27,362 | 13,045 | 170,672 |  |  |  |  |  |  | n/a | 94,442 |  |  |  |  |  |  |
| 1974 | 18,496 | 12,112 | 136,317 | 12,247 |  |  |  |  |  | n/a | 80,407 |  | ${ }^{\text {r }} 40,000$ |  |  |  |  |
| 1975 | 25,608 | 13,748 | 169,517 | 13,950 | 76,585 |  |  |  |  | n/a | 91,258 |  |  |  |  |  |  |
| 1976 | 20,791 | 12,304 | 144,894 | 13,483 | 87,230 | 54.726 |  |  |  | n/a | 81,010 |  | 20,000 |  |  |  |  |
| 1977 | 22.640 | 12,331 | 150,995 | 10,769 | 84,312 | 62.332 | 11,627 |  |  | n/a | 92.914 |  |  |  |  |  |  |
| 1978 | 29.587 | 15,128 | 191,051 | 13.392 | 67,340 | 60,247 | 13,243 | 465 |  | 154.688 | 92.428 | 0.60 |  |  |  |  |  |
| 1979 | 25,715 | 9,702 | 144,131 | 11,447 | 83,741 | 48,120 | 12,800 | 530 | 0 | 156,638 | 87,774 | 0.56 |  |  |  |  |  |
| 1980 | 16,215 | 10,927 | 121,469 | 11,929 | 71,578 | 59,840 | 10,224 | 512 | 0 | 154,082 | 57,749 | 0.37 |  | 50,000 | 100,000 <br> Egg/1000 | 150,000 | 200,000 |
| 1981 | 17.753 | 7,529 | 104,790 | 15,093 | 74,592 | 51.148 | 12.714 | 409 | 0 | 153,955 | 60,743 | 0.39 |  |  |  |  |  |


| 1981 | 17,753 | 7,529 | 104,790 | 15,093 | 74,592 | 51,148 | 12,714 | 409 | 0 | 153,955 | 60,743 | 0.39 |
| :---: | :---: | :---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :--- |
| 1982 | 13,105 | 7,472 | 89,527 | 11,386 | 94,379 | 53,301 | 10,867 | 509 | 0 | 170,443 | 59,890 | 0.35 |
| 1983 | 21,441 | 10,596 | 136,119 | 9,596 | 71,201 | 67,441 | 11,325 | 435 | 0 | 159,997 | 69,979 | 0.44 |
| 1984 | 14,792 | 8,104 | 98,957 | 8,278 | 60,005 | 50,878 | 14,329 | 453 | 0 | 133,944 | 52,401 | 0.39 |
| 1985 | 18,361 | 6,914 | 102,828 | 7,073 | 51,766 | 42,878 | 10,810 | 573 | 0 | 113,100 | 67,167 | 0.59 |
| 1986 | 27,789 | 9,300 | 148,226 | 10,753 | 44,226 | 36,991 | 9,110 | 432 | 0 | 101,513 | 77,415 | 0.76 |
| 1987 | 21,601 | 5,857 | 106,496 | 7,818 | 67,243 | 31,603 | 7,859 | 364 | 0 | 114,887 | 61,933 | 0.54 |
| 1988 | 41,829 | 5,168 | 166,964 | 8,123 | 48,885 | 48,050 | 6,715 | 314 | 0 | 112,087 | 104,517 | 0.93 |
| 1989 | 23,160 | 4,961 | 105,793 | 11,710 | 50,797 | 34,932 | 10,209 | 269 | 0 | 107,916 | 64,718 | 0.60 |
| 1990 | 22,495 | 5,048 | 104,218 | 8,413 | 73,223 | 36,298 | 7,422 | 408 | 0 | 125,765 | 62,920 | 0.50 |
| 1991 | 22,697 | 5,031 | 104,754 | 13,190 | 52,609 | 52,324 | 7,712 | 297 | 0 | 126,132 | 67,445 | 0.53 |
| 1992 | 25,634 | 6,496 | 123,490 | 8,358 | 82,480 | 37,593 | 11,117 | 308 | 0 | 139,856 | 67,000 | 0.48 |
| 1993 | 25,229 | 4,199 | 107,587 | 8,233 | 52,262 | 58,938 | 7,987 | 445 | 0 | 127,865 | 70,301 | 0.55 |
| 1994 | 21,886 | 5,685 | 106,315 | 8,276 | 51,484 | 37,345 | 12,522 | 319 | 0 | 109,946 | 61,186 | 0.56 |
| 1995 | 26,802 | 4,968 | 117,516 | 9,756 | 51,749 | 36,789 | 7,934 | 501 | 0 | 106,728 | 68,774 | 0.64 |
| 1996 | 22,483 | 3,902 | 96,889 | 8,499 | 61,004 | 36,978 | 7,816 | 317 | 0 | 114,615 | 59,207 | 0.52 |
| 1997 | 16,199 | 3,843 | 76,373 | 8,399 | 53,148 | 43,592 | 7,857 | 313 | 0 | 113,308 | 42,242 | 0.37 |
| 1998 | 22,398 | 2,583 | 88,236 | 9,284 | 52,520 | 37,978 | 9,262 | 314 | 0 | 109,358 | 59,821 | 0.55 |
| 1999 | 17,505 | 4,009 | 81,617 | 7,654 | 58,053 | 37,529 | 8,069 | 370 | 0 | 111,676 | 42,164 | 0.38 |
| 2000 | 12,546 | 1,643 | 50,670 | 6,033 | 47,863 | 41,483 | 7,974 | 323 | 0 | 103,677 | 26,753 | 0.26 |
| 2001 | 0 | 0 | 0 | 6,971 | 37,728 | 34,202 | 8,814 | 319 | 0 | 88,034 | 0 | 0.00 |
| 2002 | 0 | 0 | 0 | 6,448 | 43,589 | 26,960 | 7,267 | 353 | 0 | 84,615 | 0 | 0.00 |
| 2003 | 0 | 0 | 0 | 0 | 4,003 | 40,319 | 31,147 | 5,728 | 291 | 0 | 81,488 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 25,031 | 28,811 | 6,618 | 229 | 0 | 60,688 | 0 | 0.00 |




| Conservation limit: |  |  |  |
| :---: | :---: | :---: | :---: |
| Eggs $/ 1000$ | 1 SW | MSW | Total salmon |
| 101,513 | 21,258 | 4,356 | 25,614 |

Appendix 8c Lagged egg deposition analysis and estimation of conservation limit options - ICELAND-2

|  | Est. 1SW spawners | Est MSW spawners | Egg deposition egg $\times 10^{-3}$ | Smolt age composition |  |  |  |  |  | $\begin{array}{\|c} \text { Lagged } \\ \text { egg dep. } \\ S \\ \text { egg } \times 10^{-3} \\ \hline \end{array}$ | Total 1SW recruits R | R/S |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Egg | 5808 | 11776 |  | 1 yr | 2 yr | 3 yr | 4 yr | 5 yr | 6 yr |  |  |  |
| Fem | 25\% | 66\% |  | 0.05 | 0.55 | 0.36 | 0.06 | 0.01 | 0.00 |  |  |  |
| 1971 | 4.967 | 4.283 | 40,498 |  |  |  |  |  |  | n/a | 31.029 |  |
| 1972 | 4.519 | 6.752 | 59,040 |  |  |  |  |  |  | n/a | 34,822 |  |
| 1973 | 6,083 | 8,346 | 73,702 |  |  |  |  |  |  | n/a | 30,777 |  |
| 1974 | 5,074 | 5.912 | 53,314 | 1,903 |  |  |  |  |  | n/a | 31,638 |  |
| 1975 | 6.619 | 6.915 | 63,358 | 2.775 | 22,233 |  |  |  |  | n/a | 30,739 |  |
| 1976 | 6.098 | 5.515 | 51,716 | 3.464 | 32,413 | 14,660 |  |  |  | n/a | 37.553 |  |
| 1977 | 9,826 | 8,268 | 78,524 | 2,506 | 40,463 | 21,373 | 2,430 |  |  | $\mathrm{n} / \mathrm{a}$ | 52,346 |  |
| 1978 | 9,606 | 10,605 | 96,371 | 2,978 | 29,269 | 26,680 | 3,542 | 445 |  | 62,915 | 41,120 | 0.65 |
| 1979 | 9,241 | 6,848 | 66,640 | 2,431 | 34,784 | 19,300 | 4,422 | 649 | 0 | 61,586 | 49,219 | 0.80 |
| 1980 | 1,665 | 9,790 | 78,504 | 3,691 | 28,392 | 22,936 | 3,199 | 811 | 0 | 59,028 | 14,771 | 0.25 |
| 1981 | 7,569 | 3,466 | 37,926 | 4,529 | 43,110 | 18,721 | 3,801 | 586 | 0 | 70,748 | 29,155 | 0.41 |
| 1982 | 3.549 | 4.094 | 36,969 | 3.132 | 52,908 | 28.426 | 3.103 | 697 | 0 | 88.265 | 17.009 | 0.19 |
| 1983 | 4.818 | 2,940 | 29,847 | 3,690 | 36,585 | 34,886 | 4.711 | 569 | 0 | 80,442 | 21,981 | 0.27 |
| 1984 | 1.824 | 3.823 | 32,364 | 1.783 | 43,099 | 24,124 | 5.782 | 864 | 0 | 75,651 | 11,372 | 0.15 |
| 1985 | 10,679 | 2.373 | 33,948 | 1,738 | 20,821 | 28.418 | 3.998 | 1.060 | 0 | 56,036 | 44,325 | 0.79 |
| 1986 | 14.277 | 7.119 | 76,060 | 1.403 | 20,296 | 13,729 | 4.710 | 733 | 0 | 40,871 | 50,987 | 1.25 |
| 1987 | 8.598 | 6.750 | 64,949 | 1.521 | 16.386 | 13,383 | 2.276 | 864 | 0 | 34.429 | 32,007 | 0.93 |
| 1988 | 13,397 | 4.472 | 54,207 | 1,596 | 17.768 | 10,805 | 2.218 | 417 | 0 | 32,804 | 40,481 | 1.23 |
| 1989 | 6,979 | 3.863 | 40,155 | 3.575 | 18.637 | 11.716 | 1.791 | 407 | 0 | 36,125 | 27.248 | 0.75 |
| 1990 | 5,680 | 4.062 | 39,820 | 3.053 | 41,757 | 12.289 | 1.942 | 328 | 0 | 59,369 | 21.595 | 0.36 |
| 1991 | 9,087 | 3,178 | 37,894 | 2,548 | 35,657 | 27,534 | 2,037 | 356 | 0 | 68,131 | 32,580 | 0.48 |
| 1992 | 14.623 | 4.484 | 56,084 | 1.887 | 29,760 | 23,512 | 4.564 | 373 | 0 | 60,096 | 45,340 | 0.75 |
| 1993 | 12.027 | 4.813 | 54,867 | 1.872 | 22,045 | 19,623 | 3.897 | 837 | 0 | 48.273 | 37.598 | 0.78 |
| 1994 | 3,931 | 4,101 | 37,581 | 1,781 | 21,861 | 14,536 | 3,252 | 714 | 0 | 42,145 | 16,022 | 0.38 |
| 1995 | 9,972 | 2,591 | 34,615 | 2,636 | 20,804 | 14,415 | 2,409 | 596 | 0 | 40,860 | 31,305 | 0.77 |
| 1996 | 5,294 | 3,386 | 34,007 | 2,579 | 30,790 | 13,717 | 2,389 | 442 | 0 | 49,917 | 17,291 | 0.35 |
| 1997 | 7.700 | 2,041 | 27,043 | 1,766 | 30,122 | 20,302 | 2.274 | 438 | 0 | 54,902 | 25,181 | 0.46 |
| 1998 | 13,344 | 2,970 | 42,461 | 1,627 | 20,632 | 19,862 | 3,365 | 417 | 0 | 45,903 | 43,078 | 0.94 |
| 1999 | 6.219 | 4.929 | 47,339 | 1.598 | 19,004 | 13,604 | 3.292 | 617 | 0 | 38,115 | 18,626 | 0.49 |
| 2000 | 6,414 | 1,828 | 23,524 | 1,271 | 18,670 | 12,531 | 2,255 | 604 | 0 | 35,330 | 13,679 | 0.39 |
| 2001 | 0 | 0 | 0 | 1,996 | 14,847 | 12,311 | 2,077 | 413 | 0 | 31,643 | 0 | 0.00 |
| 2002 | 0 | 0 | 0 | 2.225 | 23,311 | 9,790 | 2.040 | 381 | 0 | 37,747 | 0 | 0.00 |
| 2003 | 0 | 0 | 0 | 1,106 | 25,989 | 15,371 | 1.623 | 374 | 0 | 44.463 | 0 | 0.00 |
| 2004 | 0 | 0 | 0 | 0 | 12,915 | 17,137 | 2,548 | 297 | 0 | 32,897 | 0 | 0.00 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 8.516 | 2,840 | 467 | 0 | 11,823 | 0 | 0.00 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 1.411 | 521 | 0 | 1,932 | 0 | 0.00 |




Appendix 8d Lagged egg deposition analysis and estimation of conservation limit options - All IRELAND

|  | Est. 1SW spawners | Est MSW spawners | $\begin{gathered} \text { Egg } \\ \text { deposition } \\ \text { egg } \times 10^{-3} \end{gathered}$ | Smolt age composition |  |  |  |  |  | $\begin{array}{\|c} \text { Lagged } \\ \text { egg dep. } \\ \mathrm{S} \\ \text { egg } \times 10^{-3} \\ \hline \end{array}$ | Total 1SW recruits R | R/S |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Egg | 3400 | 7000 |  | 1 yr | 2 yr | 3 yr | 4 yr | 5 yr | 6 yr |  |  |  |
| Fem | 60\% | 85\% |  | 0.20 | 0.70 | 0.10 | 0.00 | 0.00 | 0.00 |  |  |  |
| 1971 | 302.707 | 85,589 | 1,126,774 |  |  |  |  |  |  | n/a | 1,473,370 |  |
| 1972 | 325,211 | 95,038 | 1,228,905 |  |  |  |  |  |  | n/a | 1,570,253 |  |
| 1973 | 351,827 | 101,497 | 1,321,633 |  |  |  |  |  |  | n/a | 1,688,675 |  |
| 1974 | 385.428 | 112,261 | 1,454,228 | 225,355 |  |  |  |  |  | n/a | 1,832,924 |  |
| 1975 | 399,322 | 117.075 | 1,511,211 | 245,781 | 788,742 |  |  |  |  | n/a | 1,811,090 |  |
| 1976 | 280,702 | 82,403 | 1,062,927 | 264.327 | 860.233 | 112.677 |  |  |  | 1,237.237 | 1,304,323 | 1.05 |
| 1977 | 250,728 | 72,311 | 941,739 | 290,846 | 925,143 | 122,890 | 0 |  |  | 1,338,879 | 1,163,026 | 0.87 |
| 1978 | 221,977 | 64,593 | 837,160 | 302,242 | 1,017,959 | 132,163 | 0 | 0 |  | 1,452,365 | 1,030,334 | 0.71 |
| 1979 | 198,718 | 57,364 | 746,703 | 212.585 | 1,057,848 | 145,423 | 0 | 0 | 0 | 1,415,856 | 972,673 | 0.69 |
| 1980 | 144,168 | 64,488 | 677,809 | 188,348 | 744,049 | 151,121 | 0 | 0 | 0 | 1,083,518 | 751,420 | 0.69 |
| 1981 | 93,324 | 37,881 | 415,771 | 167,432 | 659,217 | 106,293 | 0 | 0 | 0 | 932,942 | 495,993 | 0.53 |
| 1982 | 171,802 | 16,295 | 447,432 | 149,341 | 586,012 | 94,174 | 0 | 0 | 0 | 829,527 | 879,796 | 1.06 |
| 1983 | 293,850 | 48,243 | 886,500 | 135,562 | 522,692 | 83,716 | 0 | 0 | 0 | 741.970 | 1,326,481 | 1.79 |
| 1984 | 133,372 | 62,092 | 641,526 | 83,154 | 474,466 | 74,670 | 0 | 0 | 0 | 632,291 | 691,309 | 1.09 |
| 1985 | 252,505 | 54,563 | 839,761 | 89,486 | 291,040 | 67,781 | 0 | 0 | 0 | 448,307 | 1,329,195 | 2.96 |
| 1986 | 210,026 | 79,083 | 898,994 | 177,300 | 313,202 | 41,577 | 0 | 0 | 0 | 532,079 | 1,257,963 | 2.36 |
| 1987 | 156,203 | 63,679 | 697.543 | 128,305 | 620,550 | 44,743 | 0 | 0 | 0 | 793.598 | 894,008 | 1.13 |
| 1988 | 231,810 | 68,853 | 882,570 | 167,952 | 449,068 | 88,650 | 0 | 0 | 0 | 705,670 | 1,253,827 | 1.78 |
| 1989 | 153,411 | 38,875 | 544,263 | 179,799 | 587,833 | 64,153 | 0 | 0 | 0 | 831,784 | 725,099 | 0.87 |
| 1990 | 147,612 | 30,729 | 483,967 | 139,509 | 629,296 | 83,976 | 0 | 0 | 0 | 852,781 | 531,434 | 0.62 |
| 1991 | 107,278 | 22,321 | 351,655 | 176,514 | 488,280 | 89,899 | 0 | 0 | 0 | 754,693 | 425,903 | 0.56 |
| 1992 | 180,065 | 37,713 | 591,727 | 108,853 | 617.799 | 69,754 | 0 | 0 | 0 | 796.406 | 597.981 | 0.75 |
| 1993 | 100,868 | 20,414 | 327,233 | 96,793 | 380,984 | 88,257 | 0 | 0 | 0 | 566,035 | 476,305 | 0.84 |
| 1994 | 137,626 | 45,581 | 551,962 | 70,331 | 338,777 | 54,426 | 0 | 0 | 0 | 463,534 | 633,216 | 1.37 |
| 1995 | 128,585 | 45,258 | 531,598 | 118,345 | 246,158 | 48,397 | 0 | 0 | 0 | 412,901 | 599,599 | 1.45 |
| 1996 | 121,123 | 28,023 | 413,829 | 65,447 | 414,209 | 35,165 | 0 | 0 | 0 | 514,821 | 510,517 | 0.99 |
| 1997 | 96,019 | 16,374 | 293,303 | 110,392 | 229,063 | 59,173 | 0 | 0 | 0 | 398,628 | 403,938 | 1.01 |
| 1998 | 112,674 | 21,658 | 358,724 | 106,320 | 386,373 | 32,723 | 0 | 0 | 0 | 525,416 | 441,166 | 0.84 |
| 1999 | 88,681 | 15,822 | 275,051 | 82,766 | 372,118 | 55,196 | 0 | 0 | 0 | 510,080 | 357,119 | 0.70 |
| 2000 | 113,004 | 20,131 | 350,306 | 58,661 | 289,680 | 53,160 | 0 | 0 | 0 | 401,500 | 392,298 | 0.98 |
| 2001 | 0 | 0 | 0 | 71,745 | 205,312 | 41,383 | 0 | 0 | 0 | 318,440 | 0 | 0.00 |
| 2002 | 0 | 0 | 0 | 55,010 | 251,107 | 29,330 | 0 | 0 | 0 | 335,447 | 0 | 0.00 |
| 2003 | 0 | 0 | 0 | 70,061 | 192,535 | 35,872 | 0 | 0 | 0 | 298,469 | 0 | 0.00 |
| 2004 | 0 | 0 | 0 | 0 | 245,214 | 27,505 | 0 | 0 | 0 | 272,719 | 0 | 0.00 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 35,031 | 0 | 0 | 0 | 35,031 | 0 | 0.00 |



| CL setting |  |  |  |
| :---: | :---: | :---: | :---: |
| 4.E+12 |  |  |  |
| $4 . \mathrm{E}+12 \square$ |  |  |  |
| $3 . E+12 \square$ |  |  |  |
|  |  |  |  |
|  |  |  |  |
| $\stackrel{\text { ºn }}{ }$ |  |  |  |
| $\begin{aligned} & \bar{⿻} \\ & \mathbf{X} \\ & 2 . E+12 \end{aligned}$ |  |  |  |
|  |  |  |  |
| $5 . \mathrm{E}+11$ |  |  |  |
|  |  |  |  |
| $0 . \mathrm{E}+00$ |  |  |  |
| Egg/1000 |  |  |  |
|  |  |  |  |
| Conservation limit: |  |  |  |
|  |  |  |  |
| Eggs /1000 | 1SW | MSW | Total salmon |
| 639,041 | 187,338 | 43,172 | 230,509 |

Appendix 8 e Lagged egg deposition analysis and estimation of conservation limit options - NORWAY-Combined estimates


Appendix 8 f Lagged egg deposition analysis and estimation of conservation limit options - RUSSIA

|  | Est. 1SW spawners | Est MSW spawners | Egg deposition egg $\times 10^{-3}$ | Smolt age composition |  |  |  |  |  | $\begin{array}{\|c} \text { Lagged } \\ \text { egg dep. } \\ \mathrm{S} \\ \mathrm{egg} \times 10^{-3} \\ \hline \end{array}$ | Total 1SW recruits R | R/S |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Egg | 4500 | 10500 |  | 1 yr | 2 yr | 3 yr | 4 yr | 5 yr | 6 yr |  |  |  |
| Fem | 45\% | 80\% |  | 0.00 | 0.10 | 0.70 | 0.20 | 0.00 | 0.00 |  |  |  |
| 1971 | 62,664 | 103.303 | 994.638 |  |  |  |  |  |  | n/a | 359.283 |  |
| 1972 | 68.357 | 86,456 | 864.652 |  |  |  |  |  |  | n/a | 510,861 |  |
| 1973 | 115,174 | 144,145 | 1,444,044 |  |  |  |  |  |  | n/a | 577,404 |  |
| 1974 | 105,098 | 133,315 | 1,332,668 | 0 |  |  |  |  |  | n/a | 640,415 |  |
| 1975 | 112,326 | 167.537 | 1,634,772 | 0 | 99,464 |  |  |  |  | n/a | 591,223 |  |
| 1976 | 84,457 | 143,134 | 1,373,355 | 0 | 86,465 | 696,246 |  |  |  | n/a | 443,811 |  |
| 1977 | 71,963 | 107.395 | 1,047,847 | 0 | 144,404 | 605,256 | 198.928 |  |  | 948,588 | 334.401 | 0.35 |
| 1978 | 77,491 | 73,555 | 774,785 | 0 | 133,267 | 1,010,831 | 172,930 | 0 |  | 1,317,028 | 371,200 | 0.28 |
| 1979 | 89,202 | 81.104 | 861.908 | 0 | 163,477 | 932.868 | 288,809 | 0 | 0 | 1,385,154 | 516.834 | 0.37 |
| 1980 | 58,791 | 123,751 | 1,158,558 | 0 | 137,336 | 1,144,341 | 266,534 | 0 | 0 | 1,548,210 | 343,541 | 0.22 |
| 1981 | 42,648 | 67,750 | 655,460 | 0 | 104,785 | 961,349 | 326,954 | 0 | 0 | 1,393,088 | 306,874 | 0.22 |
| 1982 | 77,861 | 81,640 | 843,445 | 0 | 77,478 | 733,493 | 274,671 | 0 | 0 | 1,085,642 | 440,480 | 0.41 |
| 1983 | 112,832 | 131.593 | 1,333,865 | 0 | 86,191 | 542,349 | 209,569 | 0 | 0 | 838.110 | 506.854 | 0.60 |
| 1984 | 106,996 | 141,157 | 1,402,384 | 0 | 115,856 | 603,336 | 154,957 | 0 | 0 | 874,149 | 527,062 | 0.60 |
| 1985 | 140,906 | 156.615 | 1,600,901 | 0 | 65,546 | 810.991 | 172.382 | 0 | 0 | 1,048,918 | 624.926 | 0.60 |
| 1986 | 123,248 | 174,572 | 1,715,980 | 0 | 84,344 | 458,822 | 231,712 | 0 | 0 | 774,878 | 412,057 | 0.53 |
| 1987 | 206,629 | 81,709 | 1,104,782 | 0 | 133,386 | 590,411 | 131,092 | 0 | 0 | 854,890 | 550,474 | 0.64 |
| 1988 | 113,569 | 89,868 | 984,871 | 0 | 140,238 | 933,705 | 168,689 | 0 | 0 | 1,242,633 | 433,672 | 0.35 |
| 1989 | 166,498 | 141,562 | 1,526,284 | 0 | 160,090 | 981,669 | 266,773 | 0 | 0 | 1,408,532 | 516,470 | 0.37 |
| 1990 | 150,593 | 140,193 | 1,482,572 | 0 | 171,598 | 1,120,631 | 280,477 | 0 | 0 | 1,572,706 | 447,242 | 0.28 |
| 1991 | 162,266 | 120,994 | 1,344,938 | 0 | 110,478 | 1,201,186 | 320,180 | 0 | 0 | 1,631,845 | 417,715 | 0.26 |
| 1992 | 135,177 | 110,798 | 1,204,433 | 0 | 98,487 | 773,347 | 343,196 | 0 | 0 | 1,215,031 | 454,788 | 0.37 |
| 1993 | 111,613 | 164,447 | 1,607,370 | 0 | 152,628 | 689,410 | 220,956 | 0 | 0 | 1,062,994 | 386,103 | 0.36 |
| 1994 | 122,714 | 142,345 | 1,444,193 | 0 | 148,257 | 1,068,399 | 196,974 | 0 | 0 | 1,413,630 | 346,539 | 0.25 |
| 1995 | 109,835 | 103,469 | 1,091,554 | 0 | 134,494 | 1,037,800 | 305,257 | 0 | 0 | 1,477,551 | 415,331 | 0.28 |
| 1996 | 134,604 | 173,972 | 1,733,938 | 0 | 120,443 | 941,457 | 296,514 | 0 | 0 | 1,358,414 | 402,650 | 0.30 |
| 1997 | 125,987 | 140,456 | 1,434,957 | 0 | 160,737 | 843.103 | 268,988 | 0 | 0 | 1,272,828 | 388,905 | 0.31 |
| 1998 | 137,515 | 140,054 | 1,454,926 | 0 | 144,419 | 1,125,159 | 240,887 | 0 | 0 | 1,510,465 | 405,794 | 0.27 |
| 1999 | 94,955 | 139,284 | 1,362,269 | 0 | 109,155 | 1,010,935 | 321,474 | 0 | 0 | 1,441,565 | 408,886 | 0.28 |
| 2000 | 109,725 | 189,057 | 1,810,268 | 0 | 173,394 | 764,088 | 288,839 | 0 | 0 | 1,226,321 | 169,067 | 0.14 |
| 2001 | 0 | 0 | 0 | 0 | 143,496 | 1,213,756 | 218,311 | 0 | 0 | 1,575,563 | 0 | 0.00 |
| 2002 | 0 |  | 0 | 0 | 145,493 | 1,004,470 | 346,788 | 0 | 0 | 1,496,750 | 0 | 0.00 |
| 2003 | 0 | 0 | 0 | 0 | 136,227 | 1,018,448 | 286,991 | 0 | 0 | 1,441,666 | 0 | 0.00 |
| 2004 | 0 | 0 | 0 | 0 | 181,027 | 953,588 | 290,985 | 0 | 0 | 1,425,600 | 0 | 0.00 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 1,267,188 | 272,454 | 0 | 0 | 1,539,641 | 0 | 0.00 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 362,054 | 0 | 0 | 362.054 | 0 | 0.00 |




| Conservation limit: |  |  |  |
| :---: | :---: | :---: | :---: |
| Eggs $/ 1000$ | 1SW | MSW | Total salmon |
| 815,922 | 75,479 | 80,808 | 156,287 |

Appendix 8 g Lagged egg deposition analysis and estimation of conservation limit options - UK(Northern Ireland)-1

|  | Est. 1SW spawners | Est MSW spawners | $\begin{array}{\|c\|} \text { Egg } \\ \text { deposition } \\ \text { egg } \times 10^{-3} \end{array}$ | Smolt age composition |  |  |  |  |  | Lagged egg dep. <br> S <br> egg $\times 10^{-3}$ | Total 1SW recruits R | R/S |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Egg | 3400 | 6000 |  | 1 yr | 2 yr | 3 yr | 4 yr | 5 yr | 6 yr |  |  |  |
| Fem | 60\% | 85\% |  | 20\% | 78\% | 2\% | 0\% | 0\% | 0\% |  |  |  |
| 1971 | 25.959 | 5.410 | 80,546 |  |  |  |  |  |  | n/a | 148.818 |  |
| 1972 | 21.015 | 4.539 | 66,019 |  |  |  |  |  |  | n/a | 123.149 |  |
| 1973 | 19,914 | 3,957 | 60,806 |  |  |  |  |  |  | n/a | 115,871 |  |
| 1974 | 23,788 | 5.051 | 74,287 | 16,109 |  |  |  |  |  | n/a | 138,122 |  |
| 1975 | 16.910 | 3.560 | 52.652 | 13,204 | 62,826 |  |  |  |  | n/a | 96,870 |  |
| 1976 | 11,756 | 2,576 | 37,121 | 12,161 | 51,495 | 1.611 |  |  |  | 65,267 | 69,660 |  |
| 1977 | 12,156 | 2,466 | 37,378 | 14,857 | 47,429 | 1,320 | 0 |  |  | 63,607 | 71,624 | 1.13 |
| 1978 | 14,485 | 3,090 | 45,311 | 10,530 | 57,944 | 1,216 | 0 | 0 |  | 69,690 | 83,923 | 1.20 |
| 1979 | 11,317 | 2,416 | 35,408 | 7,424 | 41,068 | 1,486 | 0 | 0 | 0 | 49,978 | 68,410 | 1.37 |
| 1980 | 15,372 | 3,139 | 47,369 | 7,476 | 28,955 | 1,053 | 0 | 0 | 0 | 37,483 | 87,675 | 2.34 |
| 1981 | 10,852 | 2,246 | 33,594 | 9,062 | 29,155 | 742 | 0 | 0 | 0 | 38,960 | 67,936 | 1.74 |
| 1982 | 18.931 | 3.884 | 58,428 | 7.082 | 35,343 | 748 | 0 | 0 | 0 | 43,172 | 113.168 | 2.62 |
| 1983 | 25.556 | 5.455 | 79,955 | 9,474 | 27.618 | 906 | 0 | 0 | 0 | 37,998 | 143,044 | 3.76 |
| 1984 | 9,107 | 1.875 | 28,141 | 6.719 | 36,948 | 708 | 0 | 0 | 0 | 44,375 | 55,458 | 1.25 |
| 1985 | 12,298 | 2.593 | 38,311 | 11.686 | 26,204 | 947 | 0 | 0 | 0 | 38,837 | 74,227 | 1.91 |
| 1986 | 13.890 | 2.990 | 43,581 | 15,991 | 45,574 | 672 | 0 | 0 | 0 | 62,237 | 79,531 | 1.28 |
| 1987 | 10,490 | 1.440 | 28,744 | 5,628 | 62,365 | 1,169 | 0 | 0 | 0 | 69,162 | 46,233 | 0.67 |
| 1988 | 31.897 | 5.315 | 92,179 | 7.662 | 21.950 | 1,599 | 0 | 0 | 0 | 31,211 | 106,228 | 3.40 |
| 1989 | 10,929 | 2.873 | 36,946 | 8.716 | 29,883 | 563 | 0 | 0 | 0 | 39,162 | 108,571 | 2.77 |
| 1990 | 29,685 | 4.142 | 81,680 | 5,749 | 33,993 | 766 | 0 | 0 | 0 | 40,508 | 86,873 | 2.14 |
| 1991 | 14,517 | 1,807 | 38,832 | 18,436 | 22,420 | 872 | 0 | 0 | 0 | 41,728 | 52,544 | 1.26 |
| 1992 | 38.970 | 5,110 | 105,558 | 7.389 | 71,900 | 575 | 0 | 0 | 0 | 79,864 | 116.151 | 1.45 |
| 1993 | 61,047 | 16.293 | 207.630 | 16,336 | 28.818 | 1,844 | 0 | 0 | 0 | 46,997 | 118,551 | 2.52 |
| 1994 | 20,128 | 3,641 | 59,633 | 7,766 | 63,710 | 739 | 0 | 0 | 0 | 72,216 | 77,202 | 1.07 |
| 1995 | 19,803 | 2,815 | 54,758 | 21,112 | 30,289 | 1,634 | 0 | 0 | 0 | 53,034 | 70,647 | 1.33 |
| 1996 | 26,495 | 3,664 | 72,737 | 41,526 | 82,335 | 777 | 0 | 0 | 0 | 124,638 | 74,362 | 0.60 |
| 1997 | 32.276 | 4.788 | 90,264 | 11,927 | 161.952 | 2.111 | 0 | 0 | 0 | 175,989 | 97,708 | 0.56 |
| 1998 | 128,291 | 8,275 | 303,914 | 10,952 | 46,513 | 4,153 | 0 | 0 | 0 | 61,618 | 189,494 | 3.08 |
| 1999 | 14,445 | 2.869 | 44,101 | 14,547 | 42,711 | 1,193 | 0 | 0 | 0 | 58,451 | 49,070 | 0.84 |
| 2000 | 25,342 | 3,936 | 71,771 | 18,053 | 56,735 | 1,095 | 0 | 0 | 0 | 75,883 | 64,828 | 0.85 |
| 2001 | 0 | 0 | 0 | 60,783 | 70,406 | 1,455 | 0 | 0 | 0 | 132,643 | 0 | 0.00 |
| 2002 | 0 | 0 | 0 | 8,820 | 237.053 | 1,805 | 0 | 0 | 0 | 247,679 | 0 | 0.00 |
| 2003 | 0 | 0 | 0 | 14,354 | 34,399 | 6.078 | 0 | 0 | 0 | 54,831 | 0 | 0.00 |
| 2004 | 0 | 0 | 0 | 0 | 55,981 | 882 | 0 | 0 | 0 | 56,863 | 0 | 0.00 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 1.435 | 0 | 0 | 0 | 1,435 | 0 | 0.00 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#DIV/0! |




Appendix 8h Lagged egg deposition analysis and estimation of conservation limit options - UK(Northern Ireland)-FCB area-(2)

|  | Est. 1SW spawners | Est MSW spawners | Egg deposition egg $\times 10^{-3}$ | Smolt age composition |  |  |  |  |  | $\begin{gathered} \text { Lagged } \\ \text { egg dep. } \\ S \\ \text { egg } \times 10^{-3} \\ \hline \end{gathered}$ | Total 1SW recruits R | R/S | Quasi S-R Relationship$6 . E+04$$\square$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Egg | 3400 | 6000 |  | 1 yr | 2 yr | 3 yr | 4 yr | 5 yr | 6 yr |  |  |  |  |  |  |  |  |
| Fem | 60\% | 85\% |  | 20\% | 78\% | 2\% | 0\% | 0\% | 0\% |  |  |  | $5 . E+04$ |  | ^ |  |  |
| 1971 | 12.123 | 2.432 | 37.137 |  |  |  |  |  |  | n/a | 69.530 |  | 4.E+04 |  |  |  |  |
| 1972 | 11.440 | 2.427 | 35,717 |  |  |  |  |  |  | $\mathrm{n} / \mathrm{a}$ | 66,569 |  | $\stackrel{\square}{6}$ | $\Delta$ | - 4 |  |  |
| 1973 | 9,919 | 2.078 | 30,831 |  |  |  |  |  |  | n/a | 56,274 |  |  |  | - 4 |  |  |
| 1974 | 7.460 | 1.568 | 23,215 | 7.427 |  |  |  |  |  | n/a | 44.593 |  | ${ }_{\text {区 }}^{2 . E+04}$ |  |  |  |  |
| 1975 | 8,772 | 1,881 | 27,485 | 7.143 | 28.967 |  |  |  |  | n/a | 49,994 |  |  |  |  |  |  |
| 1976 | 5,759 | 1.249 | 18.119 | 6.166 | 27.859 | 743 |  |  |  | 34,768 | 34,132 |  | .E+04 |  |  |  |  |
| 1977 | 5,661 | 1.167 | 17,501 | 4.643 | 24,048 | 714 | 0 |  |  | 29,405 | 33,756 | 1.15 |  |  |  |  |  |
| 1978 | 7.876 | 1.682 | 24,645 | 5.497 | 18.107 | 617 | 0 | 0 |  | 24.221 | 45.572 | 1.88 | 0.E+00 |  |  |  |  |
| 1979 | 4.771 | 995 | 14.808 | 3.624 | 21.438 | 464 | 0 | 0 | 0 | 25.526 | 28,229 | 1.11 |  |  |  |  |  |
| 1980 | 5.335 | 1.086 | 16,424 | 3.500 | 14,133 | 550 | 0 | 0 | 0 | 18,183 | 31,392 | 1.73 |  | 0,000 | Egg/1000 |  |  |
| 1981 | 5,427 | 1,108 | 16,720 | 4,929 | 13,650 | 362 | 0 | 0 | 0 | 18,942 | 31,288 | 1.65 |  |  |  |  |  |
| 1982 | 4,658 | 964 | 14,420 | 2.962 | 19,223 | 350 | 0 | 0 | 0 | 22,535 | 28,703 | 1.27 |  |  |  |  |  |
| 1983 | 6,936 | 1.410 | 21,340 | 3.285 | 11.550 | 493 | 0 | 0 | 0 | 15.328 | 39,203 | 2.56 |  |  |  |  |  |
| 1984 | 3,662 | 760 | 11,350 | 3,344 | 12.811 | 296 | 0 | 0 | 0 | 16,451 | 21,889 | 1.33 | CL setting |  |  |  |  |
| 1985 | 4,147 | 845 | 12,768 | 2,884 | 13,041 | 328 | 0 | 0 | 0 | 16,254 | 24,584 | 1.51 | 9.E+09 |  |  |  |  |
| 1986 | 4.313 | 945 | 13,618 | 4.268 | 11.247 | 334 | 0 | 0 | 0 | 15,850 | 25,125 | 1.59 | 8.E+09 |  |  |  |  |
| 1987 | 5,699 | 764 | 15,523 | 2,270 | 16,645 | 288 | 0 | 0 | 0 | 19,204 | 22,692 | 1.18 |  |  |  |  |  |
| 1988 | 10,678 | 1.756 | 30,740 | 2.554 | 8,853 | 427 | 0 | 0 | 0 | 11,833 | 34,184 | 2.89 | 7.E+09 |  |  |  |  |
| 1989 | 2,570 | 705 | 8,841 | 2,724 | 9,959 | 227 | 0 | 0 | 0 | 12,909 | 26,843 | 2.08 | 6.E+09 |  |  |  |  |
| 1990 | 7.476 | 1.020 | 20,450 | 3.105 | 10,622 | 255 | 0 | 0 | 0 | 13,982 | 22,035 | 1.58 |  |  |  |  |  |
| 1991 | 4,395 | 555 | 11,797 | 6,148 | 12,108 | 272 | 0 | 0 | 0 | 18,529 | 15,465 | 0.83 |  |  |  |  |  |
| 1992 | 8,935 | 1.218 | 24,437 | 1,768 | 23,977 | 310 | 0 | 0 | 0 | 26,056 | 27,404 | 1.05 | $\frac{\mathrm{O}}{\mathrm{O}}$ |  |  |  |  |
| 1993 | 15,024 | 3.993 | 51,014 | 4.090 | 6.896 | 615 | 0 | 0 | 0 | 11,601 | 29,665 | 2.56 | \% |  |  |  |  |
| 1994 | 6,049 | 1,105 | 17,975 | 2,359 | 15,951 | 177 | 0 | 0 | 0 | 18,488 | 23,463 | 1.27 |  |  |  |  |  |
| 1995 | 6,910 | 1.000 | 19,197 | 4.887 | 9,202 | 409 | 0 | 0 | 0 | 14,498 | 24,771 | 1.71 | +09 |  |  |  |  |
| 1996 | 9,262 | 1,351 | 25,784 | 10,203 | 19,061 | 236 | 0 | 0 | 0 | 29,500 | 25,860 | 0.88 | 1.E+09 |  |  |  |  |
| 1997 | 8,316 | 1.273 | 23,459 | 3,595 | 39,791 | 489 | 0 | 0 | 0 | 43,874 | 25,095 | 0.57 |  |  |  |  |  |
| 1998 | 32,049 | 2,075 | 75,962 | 3,839 | 14,020 | 1,020 | 0 | 0 | 0 | 18,880 | 48,281 | 2.56 |  |  |  |  |  |
| 1999 | 6,036 | 1,141 | 18,135 | 5,157 | 14,973 | 359 | 0 | 0 | 0 | 20,490 | 19,891 | 0.97 | 0 | 10,000 | 20,000 30,000 | 40,000 | 50,000 |
| 2000 | 8,563 | 1,331 | 24,256 | 4,692 | 20,111 | 384 | 0 | 0 | 0 | 25,187 | 22,252 | 0.88 |  |  | Egg/1000 |  |  |
| 2001 | 0 | 0 | 0 | 15,192 | 18,298 | 516 | 0 | 0 | 0 | 34,006 | 0 | 0.00 |  |  |  |  |  |
| 2002 | 0 | 0 | 0 | 3,627 | 59,251 | 469 | 0 | 0 | 0 | 63,347 | 0 | 0.00 | Conservation |  |  |  |  |
| 2003 | 0 | 0 | 0 | 4.851 | 14,145 | 1,519 | 0 | 0 | 0 | 20,516 | 0 | 0.00 | Eggs /1000 | 1SW | MSW | Total | salmon |
| 2004 | 0 | 0 | 0 | 0 | 18,920 | 363 | 0 | 0 | 0 | 19,282 | 0 | 0.00 | 11,601 | 4,436 | 626 |  | 61 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 485 | 0 | 0 | 0 | 485 | 0 | 0.00 |  |  |  |  |  |

Appenchedi Lagged eggdeposition azaysis and estimation of conservetion link optons - UK(Scotand)E


Appendix 6 ，Lagged egg deposition analysis and estimation of conservaton limitophion－UK（Scotand）$M$

|  | Exilum | $\begin{aligned} & \text { Eanny } \\ & \text { gasw } \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ＝xim | trit | ＊木17 |  | 迷亲 | 2y | 3 ${ }^{4}$ |  | 51 |  |  |  |  |
| Fint | \％ | 5me |  | 74． | －${ }^{\text {aty }}$ | Wat | ＋1＊ | 4 | 1994 |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| 189 | 419，12 | 5，5x | 121，$x^{3}$ |  |  |  |  |  |  | ก ${ }^{\text {a }}$ | cezama |  |
| 18\％． |  |  |  |  |  |  |  |  |  | Nur | 5xame |  |
| 184\％ | zanemi | 133，${ }^{\text {3 }}$ |  |  |  |  |  |  |  | nus |  |  |
| 19\％4 | 3nama | 74x | －35，106 | 44isis |  |  |  |  |  | What |  |  |
| 14F5 | 34－3 ${ }^{\text {a }}$ | 14 \％ $\mathrm{y}^{2}$ |  | 2 Lag 01 | 01063 |  |  |  |  | \％ix | F9691 |  |
| 12\％ | － 514 | $7 \mathrm{LI} \leq 5$ | 1：16．4t5 | － 4.3 | Extst | －29．vit |  |  |  | 1 ， 44, | Existu |  |
| 1979 | 3ty 4 | 4 $4 \times$ | ＋18\％ 80 | 3989］ | GMISX | YY，SE | 9 |  |  | $\underline{12412}$ | $584 \times 4$ | 1． 8 |
| 1078 | 364x | 1104 | 1．3， 3 | \％ 3.3 | 64．etm | 3 4 \％ | 0 | 7 |  |  |  | 6－5 |
| 1298 | Wegeta | 7asta | 311714 | 24539 | CET $\mathrm{STT}^{\text {a }}$ | 145，56 | $\underline{\square}$ | 0 | $\checkmark$ |  | ＊2898 | C） |
| TME | 14Fux | Wixak | \％a， | 3．4 4ici | 54xat | xuper | 5 | 4 | $\square$ | ＋xix，4t | Ti，Mat | TX |
| 14x\％ | Mi4 $4 \times$ | 14 $4 \times$ |  | 3／，प｜ | CHikM | 56\％311 | 1 | 4 | 3 |  | 37x．7es | N311 |
| 1 108x | 56，14， | 74．csis | 566atis | 18， 54.3 | cay ind | 50x， 360 | 9 | 0 | \％ |  | 54．4．749 | 2．43 |
| 1439 | 3 4 y ${ }^{\text {a }}$ |  | 17TH54 | 141845 |  | $4 \times 3$ | 4 | 4 | 1 | theres | 5－4，45 | 154 |
| 11818 |  |  | 2 15314 | IE，751 | 38，54］ | 27，514 | $\square$ | 4 | \％ | 6．Whip |  | C．7 |
| 1485 | 3tayle | 7113 | 7．4xe．tay | 191．Sex | 4xayed | 2y，y $x^{4}$ | 9 | 4 | 3 | cixay | 464． 12 | Les |
| 1404． | 172 | 113 3 x 5 | bursues | 311 $\times 23$ |  | 2t1．4 | 0 | 4 | 1 | TTM， 1 | 42， 5 \％ | 14． |
| 14．87 | 2atry | Wexty | 13xask | 200．51］ | Eil mex | $24 \times 5{ }^{3}$ | 9 | $\square$ | 0 | 1，12．${ }^{2}$ | 434．54 | C．4 |
| 1134 | 3074e | 13，xu | 417ex | －1，4\％ | －4，${ }^{4}$ |  | 9 | 4 | 3 | Hixi，$x^{4}$ | 5． 5 | \％${ }^{3}$ |
| 1489 | 26\％ 70 | 14．3．36 |  |  | 641［4\％ | 34893． | 9 | 4 |  | 1，108，741 | $44 \times 18$ | 2．44 |
| 184910 | 等 ${ }^{\text {a }}$ | W7： |  | 714．3y | 4xalm | 34x $x^{4}$ | 9 | U | 4 |  | 3＋7at？ | Cr ${ }^{4}$ |
| 141 | ＋7156 | 2， 5 |  | C 4,7 | 510， 2x $^{\text {a }}$ | X $\times 14$ | 5 | 4 | \％ | ［ IIXT 4 | 3074 | ¢3 |
| 3142 | UT／44 | 15 Ex | Y $\mathrm{Cl\mid L}$ | ＜u，］14 | Cuk | 314010 | 6 | $\underline{4}$ | 3 | vixhte | 人120 | ＋3 |
| IIIE | zabeit | n17e2 | ctra | 16，yex | cra esex | 4．2．9x | $\square$ | $\underline{\square}$ | 3 |  | ，\％r，zto | C．${ }^{4}$ |
| 144 | 4 814 | Urer | Ytiver | ｜ $1 \times 1,4$ | uty | ＋ $4 \times 1$ | 4 | $\underline{4}$ | 3 | － $4 \mathrm{c} \times \mathrm{X}$ | UYY $4 \times 4$ | Q $\times$ |
| 7189 | WW？ | Wutw | \％${ }^{\text {When }}$ | － 4 H1 | WIUTH4 |  | 0 | 4 |  | － | 3Wxy |  |
| 1459 | 5，45x | 60， 5 | Sex， 5 | 17x，301 | 515．804 | 240．861 | 6 | 5 | 3 | －4ytr | 185． | CX |
| 1897 | 86， 8 a | $33^{3} \times 5$ | SX， | 189，34 | 44－x ${ }^{\text {x }}$ | 3，\％，xic | $\underline{0}$ | 5 | 3 | 348972 | 2xisis | C22 |
| 140 | － 4 | W－ | －4¢ ${ }^{\text {ate }}$ | 16.1 | 3－14＊ |  | 4 | 4 | 3 | 9x TH | 10\％， | 413 |
| 14＊ | 44 $4 \times$ | Wax |  | 14909 | 413 | ＊W 41 | \％ | $\underline{0}$ | 3 | W¢142 | 12．${ }^{\text {che }}$ | 人1 1 |
| 20 | C3，235 | 5 | 46．43 | 12． 204 | 2T210 | 3115 | $\underline{\square}$ | $\square$ | 3 | 5.8 .812 | T6． 3.31 | C12 |
| 200 | \％ | $\Gamma$ | V | ［21．89］ | 26x ${ }^{\text {8in }}$ | 170．38 | 5 | $\square$ | 7 | $5 \mathrm{El}, 7 \mathrm{~F}$ | 1 | C．EI |
| MKY | \％ | 0 | 4 | Sx，XY | Wrexu | 14， $3 \times$ | 0 | 3 | 0 | 5x，160 | $\underline{0}$ | CWI |
| ，0113 | 0 | I | 8 | 9，（17 | 60，bry | 101，42 | 4 | $\square$ | 3 | 4，＜1E5 | $\underline{\square}$ | COL |
| 3 T | 4 | D | $\underline{4}$ | 9 | 20ylm | \＄ 4 ， 2 C | $\underline{1}$ | $t$ | 0 | \＃Yrem | $\underline{\square}$ | S．19 |
| We | P | （1） | I | 4 | 1 |  | U | 4 | 通 | TJYM | 1 | Wlu |
| 2TE | ¢ | $\square$ | V | 01 | 4 | 4 | 0 | 4 | \％ | 1 | 1 | WUVM |



Appendix 8k Estimated numbers of fish killed and recruits from Monte Carlo simulation analysis FAROES

| Year | Estimated total catch 1SW |  | Estimated total catch MSW |  | Est. mat. 1SW recruits mean |  | Est. nonmat. 1SW recruits mran |  | Total 1SW recruits means |  | Prop'n wild | Stock compositio | 1SW | MSW |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 12,358 | $9.08 \mathrm{E}+06$ | 105,796 | 9.45E-08 | 2,717 | $5.27 \mathrm{E}+05$ | 122,575 | $6.22 \mathrm{E}+06$ | 125,292 | 2,598 | 1.00 | France | 0.05 | 0 |
| 1972 | 13.031 | $1.10 \mathrm{E}+07$ | 111.187 | 1.36E-07 | 2.869 | $6.51 \mathrm{E}+05$ | 138.149 | $7.32 \mathrm{E}+06$ | 141.018 | 2.823 | 1.00 | Finland | 0.05 | 0 |
| 1973 | 14,802 | $1.37 \mathrm{E}+07$ | 126,012 | $8.40 \mathrm{E}-08$ | 3,263 | $8.53 \mathrm{E}+05$ | 101,227 | $9.06 \mathrm{E}+06$ | 104,490 | 3,148 | 1.00 | Iceland | 0 | 0.006 |
| 1974 | 10,331 | $6.76 \mathrm{E}+06$ | 88.276 | $7.88 \mathrm{E}-08$ | 2.278 | $4.20 \mathrm{E}+05$ | 122.803 | $4.53 \mathrm{E}+06$ | 125,080 | 2.226 | 1.00 | Ireland | 0.1 | 0.057 |
| 1975 | 13,210 | $1.07 \mathrm{E}+07$ | 112,984 | $5.06 \mathrm{E}-08$ | 2,908 | $6.39 \mathrm{E}+05$ | 85,387 | $7.01 \mathrm{E}+06$ | 88,296 | 2,766 | 1.00 | Norway | 0.3 | 0.396 |
| 1976 | 8.649 | $4.86 \mathrm{E}+06$ | 73.900 | $2.34 \mathrm{E}-08$ | 1.903 | $2.84 \mathrm{E}+05$ | 59,691 | $3.17 \mathrm{E}+06$ | 61.593 | 1.859 | 1.00 | Russia | 0.1 | 0.183 |
| 1977 | 6.110 | $2.23 \mathrm{E}+06$ | 52.112 | $6.21 \mathrm{E}-09$ | 1.342 | $1.28 \mathrm{E}+05$ | 44.702 | $1.47 \mathrm{E}+06$ | 46.045 | 1.264 | 1.00 | Sweden | 0.05 | 0.023 |
| 1978 | 4,602 | $1.35 \mathrm{E}+06$ | 39,309 | $1.07 \mathrm{E}-09$ | 1,014 | 8.12E+04 | 74,756 | $9.72 \mathrm{E}+05$ | 75,770 | 1,026 | 1.00 | UK(E\&W) | 0.1 | 0.023 |
| 1979 | 8.152 | $4.41 \mathrm{E}+06$ | 70,082 | $4.77 \mathrm{E}-10$ | 1.796 | $2.69 \mathrm{E}+05$ | 191.784 | $3.82 \mathrm{E}+06$ | 193.580 | 2.022 | 1.00 | UK(NI) | 0.05 | 0 |
| 1980 | 21.067 | $2.67 \mathrm{E}+07$ | 182.617 | $1.20 \mathrm{E}-07$ | 4.652 | $1.75 \mathrm{E}+06$ | 321.621 | $1.87 \mathrm{E}+07$ | 326.273 | 4.524 | 1.00 | UK(Sc) | 0.2 | 0.192 |
| 1981 | 35,317 | $7.72 \mathrm{E}+07$ | 300,542 | $5.58 \mathrm{E}-07$ | 7,778 | $4.73 \mathrm{E}+06$ | 308,871 | $5.08 \mathrm{E}+07$ | 316,649 | 7,453 | 0.98 |  |  |  |
| 1982 | 32.485 | $6.57 \mathrm{E}+07$ | 276.957 | 6.49E-07 | 7.168 | $4.14 \mathrm{E}+06$ | 244,092 | $4.25 \mathrm{E}+07$ | 251.260 | 6.828 | 0.98 | Other |  | 0.122 |
| 1983 | 36,381 | $4.51 \mathrm{E}+07$ | 215,350 | $2.41 \mathrm{E}-07$ | 8,020 | $3.28 \mathrm{E}+06$ | 168,863 | $2.96 \mathrm{E}+07$ | 176.883 | 5.737 | 0.98 |  |  |  |
| 1984 | 18,528 | $1.69 \mathrm{E}+07$ | 138,227 | 8.97E-08 | 4,083 | $1.09 \mathrm{E}+06$ | 175,039 | $1.15 \mathrm{E}+07$ | 179,122 | 3,545 | 0.96 | Total | 1 | 1.002 |
| 1985 | 14,816 | $2.04 \mathrm{E}+07$ | 158,103 | 2.33E-07 | 3,264 | $1.17 \mathrm{E}+06$ | 195,303 | $1.35 \mathrm{E}+07$ | 198,567 | 3,834 | 0.92 |  |  |  |
| 1986 | 18,206 | $2.58 \mathrm{E}+07$ | 180,934 | $2.16 \mathrm{E}-07$ | 4.012 | $1.52 \mathrm{E}+06$ | 183.050 | $1.69 \mathrm{E}+07$ | 187.061 | 4.293 | 0.96 |  |  |  |
| 1987 | 15.123 | $2.28 \mathrm{E}+07$ | 166.244 | 1.97E-07 | 3.334 | $1.29 \mathrm{E}+06$ | 100,822 | $1.45 \mathrm{E}+07$ | 104.155 | 3.973 | 0.97 |  |  |  |
| 1988 | 17,333 | $7.30 \mathrm{E}+06$ | 87,629 | $4.92 \mathrm{E}-08$ | 3,822 | $6.00 \mathrm{E}+05$ | 137,412 | $5.07 \mathrm{E}+06$ | 141,233 | 2,381 | 0.92 |  |  |  |
| 1989 | 12.865 | $1.31 \mathrm{E}+07$ | 121,965 | $7.73 \mathrm{E}-08$ | 2.834 | $7.68 \mathrm{E}+05$ | 152,273 | $8.63 \mathrm{E}+06$ | 155,107 | 3.065 | 0.82 |  |  |  |
| 1990 | 14.836 | $1.65 \mathrm{E}+07$ | 140,054 | $1.05 \mathrm{E}-07$ | 3.270 | $9.95 \mathrm{E}+05$ | 97.862 | $1.06 \mathrm{E}+07$ | 101,132 | 3.409 | 0.54 |  |  |  |
| 1991 | 8,584 | $6.09 \mathrm{E}+06$ | 84,935 | $4.68 \mathrm{E}-08$ | 1,892 | $3.52 \mathrm{E}+05$ | 42,975 | $3.83 \mathrm{E}+06$ | 44,867 | 2,046 | 0.54 |  |  |  |
| 1992 | 3.316 | $1.05 \mathrm{E}+06$ | 35.700 | $7.04 \mathrm{E}-09$ | 732 | $6.16 \mathrm{E}+04$ | 33,079 | $6.81 \mathrm{E}+05$ | 33,811 | 862 | 0.62 |  |  |  |
| 1993 | 2.690 | $7.29 \mathrm{E}+05$ | 30,023 | $2.45 \mathrm{E}-09$ | 594 | $4.23 \mathrm{E}+04$ | 34.262 | $5.03 \mathrm{E}+05$ | 34.856 | 738 | 0.69 |  |  |  |
| 1994 | 2,863 | $8.59 \mathrm{E}+05$ | 31,672 | $5.85 \mathrm{E}-09$ | 631 | $4.81 \mathrm{E}+04$ | 37,434 | $5.75 \mathrm{E}+05$ | 38,065 | 790 | 0.72 |  |  |  |
| 1995 | 3.130 | $1.00 \mathrm{E}+06$ | 34.662 | $1.04 \mathrm{E}-08$ | 688 | $5.45 \mathrm{E}+04$ | 31,269 | $6.45 \mathrm{E}+05$ | 31,957 | 837 | 0.80 |  |  |  |
| 1996 | 2.639 | $6.33 \mathrm{E}+05$ | 28.381 | $6.03 \mathrm{E}-09$ | 581 | $3.60 \mathrm{E}+04$ | 2.071 | $3.98 \mathrm{E}+05$ | 2.652 | 659 | 0.75 |  |  |  |
| 1997 | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 1.446 | $5.58 \mathrm{E}+01$ | 1.446 | 7 | 0.80 |  |  |  |
| 1998 | 590 | $3.13 \mathrm{E}+03$ | 1,424 | $1.75 \mathrm{E}-11$ | 130 | $4.38 \mathrm{E}+02$ | 463 | $2.18 \mathrm{E}+03$ | 593 | 51 | 0.80 |  |  |  |
| 1999 | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 1.792 | $8.57 \mathrm{E}+01$ | 1.792 | 9 | 0.80 |  |  |  |
| 2000 | 510 | $4.19 \mathrm{E}+03$ | 1,765 | $3.10 \mathrm{E}-11$ | 112 | $4.20 \mathrm{E}+02$ | 400 | $2.77 \mathrm{E}+03$ | 513 | 56 | 0.80 |  |  |  |
| 2001 | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | 0 | 0.80 |  |  |  |
| 2002 | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | 0 | 0.80 |  |  |  |
| 2003 | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | 0 | 0.80 |  |  |  |
| 2004 | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | 0 | 0.80 |  |  |  |
| 2005 | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | 0 | 0.80 |  |  |  |

Appendix 81 Estimated numbers of fish killed and recruits from Monte Carlo simulation analysis - WEST GREENLAND

| Year | Estimated totalcatch 1SW | $\checkmark$ ariance | Estim ated totalcatch M S W | $V$ ariance | Estimated number 1 SW recruits mean | Variance | Est. non-mat. 1SW recruits mean | $\checkmark$ ariance | Prop'n EU | $\begin{array}{r} \text { European } \\ \text { stock } \\ \text { composition } \end{array}$ | M S W |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 0 | $0.00 \mathrm{E}+00$ | 951.713 | $9.69 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 1.040 .849 | $1.89 \mathrm{E}+09$ | 0.50 | France | 0.027 |
| 1972 | 0 | $0.00 \mathrm{E}+00$ | 683.518 | $4.91 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 747.549 | $9.89 \mathrm{E}+08$ | 0.50 | Finland | 0.001 |
| 1973 | 0 | $0.00 \mathrm{E}+00$ | 623.043 | $4.02 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 681.415 | $8.21 \mathrm{E}+08$ | 0.50 | Iceland | 0.001 |
| 1974 | 0 | $0.00 \mathrm{E}+00$ | 595.929 | $3.70 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 651.746 | $7.34 \mathrm{E}+08$ | 0.50 | Ireland | 0.147 |
| 1975 | 0 | $0.00 \mathrm{E}+00$ | 723.169 | $5.53 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 790.896 | $1.08 \mathrm{E}+09$ | 0.50 | Norway | 0.027 |
| 1976 | 0 | $0.00 \mathrm{E}+00$ | 429.572 | $1.92 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 469.788 | $3.63 \mathrm{E}+08$ | 0.50 | Russia | 0.000 |
| 1977 | 0 | $0.00 \mathrm{E}+00$ | 492.205 | $2.52 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 538.302 | $4.96 \mathrm{E}+08$ | 0.50 | Sweden | 0.003 |
| 1978 | 0 | $0.00 \mathrm{E}+00$ | 326.682 | $1.04 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 357.281 | $2.14 \mathrm{E}+08$ | 0.48 | UK(E\&W) | 0.149 |
| 1979 | 0 | $0.00 \mathrm{E}+00$ | 463.787 | $2.24 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 507.238 | $4.57 \mathrm{E}+08$ | 0.50 | UK(NI) | 0.000 |
| 1980 | 0 | $0.00 \mathrm{E}+00$ | 412.993 | $1.77 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 451.702 | $3.77 \mathrm{E}+08$ | 0.52 | UK(Sc) | 0.645 |
| 1981 | 0 | $0.00 \mathrm{E}+00$ | 443.665 | $2.01 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 485.228 | $4.10 \mathrm{E}+08$ | 0.41 |  |  |
| 1982 | 0 | $0.00 \mathrm{E}+00$ | 385.618 | $1.54 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 421.732 | $3.03 \mathrm{E}+08$ | 0.38 | Other |  |
| 1983 | 0 | $0.00 \mathrm{E}+00$ | 111.061 | $1.32 \mathrm{E}+07$ | 0 | $0.00 \mathrm{E}+00$ | 121.468 | $2.71 \mathrm{E}+07$ | 0.60 |  |  |
| 1984 | 0 | $0.00 \mathrm{E}+00$ | 106,146 | $1.15 \mathrm{E}+07$ | 0 | $0.00 \mathrm{E}+00$ | 116,090 | $2.36 \mathrm{E}+07$ | 0.50 | Total | 1.000 |
| 1985 | 0 | $0.00 \mathrm{E}+00$ | 334.696 | $1.18 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 366,061 | $2.46 \mathrm{E}+08$ | 0.50 |  |  |
| 1986 | 0 | $0.00 \mathrm{E}+00$ | 351.775 | $1.28 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 384.733 | $2.63 \mathrm{E}+08$ | 0.43 |  |  |
| 1987 | 0 | $0.00 \mathrm{E}+00$ | 340.041 | $1.18 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 371.884 | $2.31 \mathrm{E}+08$ | 0.41 |  |  |
| 1988 | 0 | $0.00 \mathrm{E}+00$ | 312.145 | $1.06 \mathrm{E}+08$ | 0 | $0.00 \mathrm{E}+00$ | 341.383 | $2.08 \mathrm{E}+08$ | 0.57 |  |  |
| 1989 | 0 | $0.00 \mathrm{E}+00$ | 130.538 | $1.76 \mathrm{E}+07$ | 0 | $0.00 \mathrm{E}+00$ | 142.769 | $3.63 \mathrm{E}+07$ | 0.44 |  |  |
| 1990 | 0 | $0.00 \mathrm{E}+00$ | 113.195 | $1.36 \mathrm{E}+07$ | 0 | $0.00 \mathrm{E}+00$ | 123.796 | $2.66 \mathrm{E}+07$ | 0.25 |  |  |
| 1991 | 0 | $0.00 \mathrm{E}+00$ | 198.244 | $3.99 \mathrm{E}+07$ | 0 | $0.00 \mathrm{E}+00$ | 216.823 | $8.46 \mathrm{E}+07$ | 0.35 |  |  |
| 1992 | 0 | $0.00 \mathrm{E}+00$ | 94.036 | $9.31 \mathrm{E}+06$ | 0 | $0.00 \mathrm{E}+00$ | 102.846 | $1.90 \mathrm{E}+07$ | 0.46 |  |  |
| 1993 | 0 | $0.00 \mathrm{E}+00$ | 2.047 | $9.20 \mathrm{E}+04$ | 0 | $0.00 \mathrm{E}+00$ | 2.239 | $1.15 \mathrm{E}+05$ | 0.3 |  |  |
| 1994 | 0 | $0.00 \mathrm{E}+00$ | 2.032 | $9.16 \mathrm{E}+04$ | 0 | $0.00 \mathrm{E}+00$ | 2.223 | $1.12 \mathrm{E}+05$ | 0.3 |  |  |
| 1995 | 0 | $0.00 \mathrm{E}+00$ | 36.071 | $1.38 \mathrm{E}+06$ | 0 | $0.00 \mathrm{E}+00$ | 39.448 | $2.66 \mathrm{E}+06$ | 0.32 |  |  |
| 1996 | 0 | $0.00 \mathrm{E}+00$ | 37.751 | $1.72 \mathrm{E}+06$ | 0 | $0.00 \mathrm{E}+00$ | 41.288 | $3.31 \mathrm{E}+06$ | 0.27 |  |  |
| 1997 | 0 | $0.00 \mathrm{E}+00$ | 24.931 | $7.06 \mathrm{E}+05$ | 0 | $0.00 \mathrm{E}+00$ | 27.266 | $1.39 E+06$ | 0.20 |  |  |
| 1998 | 0 | $0.00 \mathrm{E}+00$ | 4.301 | $1.81 \mathrm{E}+04$ | 0 | $0.00 \mathrm{E}+00$ | 4.703 | $3.44 \mathrm{E}+04$ | 0.21 |  |  |
| 1999 | 0 | $0.00 \mathrm{E}+00$ | 12.497 | $2.13 \mathrm{E}+06$ | 0 | $0.00 \mathrm{E}+00$ | 13.668 | $2.68 \mathrm{E}+06$ | 0.10 |  |  |
| 2000 | 0 | $0.00 \mathrm{E}+00$ | 13,724 | $1.76 \mathrm{E}+06$ | 0 | $0.00 \mathrm{E}+00$ | 15,009 | $2.24 \mathrm{E}+06$ | 0.30 |  |  |
| 2001 | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0.00 |  |  |
| 2002 | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0.00 |  |  |
| 2003 | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0.00 |  |  |
| 2004 | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0.00 |  |  |
| 2005 | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0 | $0.00 \mathrm{E}+00$ | 0.00 |  |  |

Data from Spain received after the Working Group meeting but included as an appendix for information

| RIVER | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | X:1991-00 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BIDASOA | 22 | 59 | 59 | 64 | 49 | 47 | 38 | 36 | 21 | 35 | 43 | NAVARRA |
| PAS | 453 | © 6 | 16 | 77 | 36 | 106 | 30 | 12 | 19 | 7 | 84 | CANTABRIA |
| NANSA | 65 | 57 | 44 | 35 | 74 | 86 | 29 | 17 | 25 | 82 | 51 |  |
| ASON | 101 | 226 | 75 | 76 | 86 | 66 | 48 | 16 | 32 | 48 | 77 |  |
| DEVA | 2 | क6 | 20 | 5 | 23 | 209 | 46 | 8 | 7 | 95 | 48 |  |
| other | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| DEVA-CARES | 455 | 921 | 496 | 506 | 247 | 252 | 117 | 177 | 183 | 413 | 377 | ASTURIAS |
| NARCEA | 766 | 430 | 917 | 1038 | 836 | 401 | 270 | 365 | 394 | 643 | 606 |  |
| SELLA | 297 | 543 | 423 | 581 | 219 | 192 | 130 | 259 | 447 | 440 | 353 |  |
| ESVA | 204 | 169 | 331 | 201 | 175 | 142 | 55 | 66 | 64 | 102 | 151 |  |
| NAVIA | 13 | 9 | 16 | 8 | 3 | 3 | 0 | 1 | 0 | 1 | 5 |  |
| EO | 94 | 46 | 27 | 52 | 22 | 26 | 16 | 12 | 54 | 43 | 39 |  |
| other | 3 | 6 | 21 | 7 | 5 | 9 | 6 | 4 | 1 | 18 | 8 |  |
| MASMA | 84 | 32 | 24 | 22 | 5 | 13 | 3 | 2 | 2 | 6 | 19 | GALICIA |
| ตaiNo | 29 | 38 | 3 | 17 | 12 | 18 | 8 | 0 | 0 | 0 | 13 |  |
| ULLA | 33 | 21 | 1 | 4 | 1 | 2 | 1 | 0 | 0 | 0 | 6 |  |
| MANDEO | 11 | 9 | 2 | 12 | 0 | 5 | 0 | 4 | 0 | 5 | 5 |  |
| other | 3 | 7 | 16 | 9 | 0 | 2 | 0 | 2 | 0 | 0 | 5 |  |
| total catch Sp | 2636 | 2725 | 2490 | 2714 | 1793 | 1579 | 797 | 979 | 1249 | 1938 | 1890 | SPAIN |

## REPORT OF THE

# ICES COMPILATION OF MICROTAGS, FINCLIP AND EXTERNAL TAG RELEASES 2000 

# by <br> THE WORKING GROUP ON NORTH ATLANTIC SALMON 

Aberdeen, Scotland<br>2-11 April 2001

This report is not to be quoted without prior consultation with the General Secretary. The document is a report of an expert group under the auspices of the International Council for the Exploration of the Sea and does not necessarily represent the views of the Council.

International Council for the Exploration of the Sea
Conseil International pour l'Exploration de la Mer
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## TERMS OF REFERENCE

The terms of reference for the 2001 Working Group on North Atlantic Salmon (C.Res. 2000/2ACFM07) stated that the Group should
"With respect to the Atlantic salmon in the NASCO area, provide a compilation of tag releases by countries in 2000".

Data were provided by Working Group members for national tagging programs, as far as possible including all agencies and organizations. These compilations for 2000 are presented by country together with a summary of the tags and marks by all countries (Table 1). Data were supplied in the standard format agreed by the Working Group in 1997. A list of national tag clearing houses is also given (Appendix 1).

ICES COMPILATION OF MICROTAG, FIN CLIP, AND EXTERNAL TAG RELEASES
Marking Season: 2000

## Country: Canada

| Origin | Primary Tag or Mark |  |  | Total |
| :---: | :---: | :---: | :---: | :---: |
|  | Microtag | External mark | Adipose clip |  |
| Hatchery juvenile | 0 | 45,009 | 1,738,916 | 1,783,925 |
| Wild juvenile | 0 | 9,083 | 329 | 9,412 |
| Adult | 0 | 6,046 | 0 | 6,046 |
| Total | 0 | 60,138 | 1,739,245 | 1,799,383 |



[^14]1 Canada continued.

| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2+, 3+ and 4+ | Smolt | W | LSW Miramichi | STR-green | 451 | NB02301-NB02800 | none | May 5-June 1 | LSW Miramichi | smolt wheel |
|  | 2+, 3+ and 4+ | Smolt | W | LSW Miramichi | STR-green | 784 | NW19000-NW 19991 | none | May 5-June 1 | LSW Miramichi | smolt wheel |
|  | 2+, 3+ and 4+ | Smolt | W | LSW Miramichi | STR-green | 50 | NB02751-NB02800 | none | May 5-June 1 | Catamaran Brook | fyke net |
|  | 2+, 3+ and 4+ | Smolt | W | NW Miramichi | STR-green | 2834 | NW10000-NW13049 | none | May 13 - Jund | NW Miramichi | estuarial trap |
|  | 1SW | Adult | W | Miramichi | Carlin-blue | 48 | zz62850-zz62897 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | 1SW | Adult | W | Miramichi | Carlin-blue | 100 | zz65500-zz65599 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | 1SW | Adult | W | Miramichi | Carlin-blue | 200 | zz65700-zz65899 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | 1SW | Adult | W | Miramichi | Carlin-blue | 861 | zz66137-zz66999 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | MSW / 1SW | Adult | w | Miramichi | Carlin-blue | 49 | zz74800-zz74849 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | 1SW | Adult | W | Miramichi | Carlin-blue | 350 | zz80650-zz80999 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | MSW | Adult | W | Miramichi | Carlin-blue | 200 | zz82300-zz82499 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | MSW | Adult | W | Miramichi | Carlin-blue | 153 | zz82722-zz82999 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | MSW / 1SW | Adult | W | Miramichi | Carlin-blue | 282 | zz85568-zz85999 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | MSW / 1SW | Adult | W | Miramichi | Carlin-blue | 611 | zz86321-zz86999 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | MSW / 1SW | Adult | W | Miramichi | Carlin-blue | 6 | zz87993-zz87998 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | 1SW | Adult | W | Miramichi | Carlin-blue | 340 | zz90000-zz90340 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | 1SW | Adult | W | Miramichi | Carlin-blue | 936 | zz91000-zz91999 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | MSW | Adult | W | Miramichi | Carlin-blue | 106 | zz92100-zz92253 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | MSW | Adult | W | Miramichi | Carlin-blue | 2 | zz92896-zz92899 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | MSW | Adult | W | Miramichi | Carlin-blue | 100 | zz93000-zz93099 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | MSW / 1SW | Adult | W | Miramichi | Carlin-blue | 363 | zz94000-zz94649 | none | June - Oct | NW and SW Miramichi R | estuarial traps |
|  | MSW / 1SW | Adult | W | LSW Miramichi | Carlin-blue | 35 | zz52962-zz52998 | none | Oct - Nov | Catamaran Brook | counting fence |
|  | MSW / 1SW | Adult | W | LSW Miramichi | Carlin-blue | 8 | zz81240-zz81247 | none | Oct - Nov | Catamaran Brook | counting fence |
|  |  |  |  |  |  |  |  |  |  |  |  |
| NF Region |  | Kelts | W | Campbellton River | T-bar-orange | 446 | T41201-41301 | none | April/May | Campbellton R |  |
|  |  |  |  |  |  |  | T41303-T43109 | none |  |  |  |
|  |  |  |  |  |  |  | T43113-T43150 | none |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | Kelts | w | Campbellton River | I-button | 49 | 987 | none | April/May | Campbellton R | 4D7491 |
|  |  |  |  |  | Archieval \& |  | 913 | none |  |  | 5072B0 |
|  |  |  |  |  | Red Atkins |  | 989 | none |  |  | 317418 |
|  |  |  |  |  |  |  | 984 | none |  |  | 9A7219 |
|  |  |  |  |  |  |  | 922 | none |  |  | D774F2 |
|  |  |  |  |  |  |  | 995 | none |  |  | 5E72C9 |
|  |  |  |  |  |  |  | 959 | none |  |  | FD7484 |
|  |  |  |  |  |  |  | 903 | none |  |  | F374FD |
|  |  |  |  |  |  |  | 944 | none |  |  | FA752D |

1 Canada continued.

| Marking <br> Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | 904 | none |  |  | 33740B |
|  |  |  |  |  |  |  | 969 | none |  |  | C77325 |
|  |  |  |  |  |  |  | 986 | none |  |  | 1 E7382 |
|  |  |  |  |  |  |  | 942 | none |  |  | 007363 |
|  |  |  |  |  |  |  | 980 | none |  |  | 3D7472 |
|  |  |  |  |  |  |  | 979 | none |  |  | CF7426 |
|  |  |  |  |  |  |  | 946 | none |  |  | 917495 |
|  |  |  |  |  |  |  | 934 | none |  |  | 3A74C2 |
|  |  |  |  |  |  |  | 923 | none |  |  | EB7530 |
|  |  |  |  |  |  |  | 966 | none |  |  | 8F756D |
|  |  |  |  |  |  |  | 917 | none |  |  | 4173B4 |
|  |  |  |  |  |  |  | 924 | none |  |  | 61739D |
|  |  |  |  |  |  |  | 996 | none |  |  | D574E1 |
|  |  |  |  |  |  |  | 915 | none |  |  | BD74D6 |
|  |  |  |  |  |  |  | 908 | none |  |  | BB73A6 |
|  |  |  |  |  |  |  | 961 | none |  |  | 961 |
|  |  |  |  |  |  |  | 926 | none |  |  | 926 |
|  |  |  |  |  |  |  | 999 | none |  |  | 999 |
|  |  |  |  |  |  |  | 957 | none |  |  | 957 |
|  |  |  |  |  |  |  | 928 | none |  |  | 928 |
|  |  |  |  |  |  |  | 910 | none |  |  | 297383 |
|  |  |  |  |  |  |  | 953 | none |  |  | F27337 |
|  |  |  |  |  |  |  | 954 | none |  |  | 5D7409 |
|  |  |  |  |  |  |  | 960 | none |  |  | AC7384 |
|  |  |  |  |  |  |  | 992 | none |  |  | E39233 |
|  |  |  |  |  |  |  | L1904 | none |  |  | 2E747C |
|  |  |  |  |  |  |  | L1915 | none |  |  | 8D73C7 |
|  |  |  |  |  |  |  | L1934 | none |  |  | 8D7231 |
|  |  |  |  |  |  |  | L1966 | none |  |  | C97245 |
|  |  |  |  |  |  |  | L1930 | none |  |  | A872BB |
|  |  |  |  |  |  |  | L1937 | none |  |  | 577200 |
|  |  |  |  |  |  |  | L1954 | none |  |  | 42723B |
|  |  |  |  |  |  |  | L1955 | none |  |  | C2729E |
|  |  |  |  |  |  |  | L1971 | none |  |  | DF72E8 |
|  |  |  |  |  |  |  | L1969 | none |  |  | 4972E1 |
|  |  |  |  |  |  |  | L1987 | none |  |  | 0C7309 |

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1 Canada continued.

| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | L1986 | none |  |  | 2C7576 |
|  |  |  |  |  |  |  | L1960 | none |  |  | A17368 |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | Kelt | W | Rocky River | I button Archival | 1 | N3993 | ADC | May | Rocky R | 34C0000073BD |
|  |  |  |  |  | \& Green Atkins |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | Kelts | W | Western Arm Bk | I-Button Archival 8 | 97 | 809 | none | May-Jun | Western Arm Brook | 34C00000751C |
|  |  |  |  |  | Red Atkins |  | 824 | none |  |  | 34C00000753D |
|  |  |  |  |  |  |  | 823 | none |  |  | 34C00000746D |
|  |  |  |  |  |  |  | 807 | none |  |  | 34C000007526 |
|  |  |  |  |  |  |  | 805 | none |  |  | 34C00000734B |
|  |  |  |  |  |  |  | 819 | none |  |  | $34 \mathrm{C000007287}$ |
|  |  |  |  |  |  |  | 810 | none |  |  | 34C000007411 |
|  |  |  |  |  |  |  | 812 | none |  |  | 34C0000074BA |
|  |  |  |  |  |  |  | 815 | none |  |  | 34C000007416 |
|  |  |  |  |  |  |  | 811 | none |  |  | 34C00000731A |
|  |  |  |  |  |  |  | 820 | none |  |  | 34C0000072BA |
|  |  |  |  |  |  |  | 801 | none |  |  | 34C0000072EC |
|  |  |  |  |  |  |  | 802 | none |  |  | 34C0000073BC |
|  |  |  |  |  |  |  | 806 | none |  |  | 34C00000736A |
|  |  |  |  |  |  |  | 816 | none |  |  | 34C000007517 |
|  |  |  |  |  |  |  | 817 | none |  |  | 34C00000736C |
|  |  |  |  |  |  |  | 818 | none |  |  | 34C0000074B7 |
|  |  |  |  |  |  |  | 803 | none |  |  | 34C0000074F8 |
|  |  |  |  |  |  |  | 808 | none |  |  | 34C0000072A6 |
|  |  |  |  |  |  |  | 804 | none |  |  | 34C0000072CD |
|  |  |  |  |  |  |  | 825 | none |  |  | $34 \mathrm{C0000074CD}$ |
|  |  |  |  |  |  |  | 800 | none |  |  | $34 \mathrm{C0000072} \mathrm{AD}$ |
|  |  |  |  |  |  |  | 822 | none |  |  | 34C000007557 |
|  |  |  |  |  |  |  | 814 | none |  |  | 34C00000727E |
|  |  |  |  |  |  |  | 821 | none |  |  | 34C0000073D5 |
|  |  |  |  |  |  |  | 826 | none |  |  | 34C00000750A |
|  |  |  |  |  |  |  | 827 | none |  |  | 34C0000072F4 |
|  |  |  |  |  |  |  | 828 | none |  |  | 34C000007456 |
|  |  |  |  |  |  |  | 829 | none |  |  | 34C000007203 |
|  |  |  |  |  |  |  | 830 | none |  |  | 34C0000074CF |

1 Canada continued.

| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | 831 | none |  |  | 34C000007498 |
|  |  |  |  |  |  |  | 832 | none |  |  | 34C00000722B |
|  |  |  |  |  |  |  | 833 | none |  |  | 34C0000072ED |
|  |  |  |  |  |  |  | 834 | none |  |  | 34C000007386 |
|  |  |  |  |  |  |  | 835 | none |  |  | 34C000007367 |
|  |  |  |  |  |  |  | 836 | none |  |  | 34C00000732A |
|  |  |  |  |  |  |  | 837 | none |  |  | 34C00000720D |
|  |  |  |  |  |  |  | 839 | none |  |  | 34C00000735E |
|  |  |  |  |  |  |  | 840 | none |  |  | 34C00000724F |
|  |  |  |  |  |  |  | 841 | none |  |  | 34C0000072A1 |
|  |  |  |  |  |  |  | 842 | none |  |  | 34C000007376 |
|  |  |  |  |  |  |  | 843 | none |  |  | 34C0000072A9 |
|  |  |  |  |  |  |  | 844 | none |  |  | 34C0000074D8 |
|  |  |  |  |  |  |  | 845 | none |  |  | 34C0000073A0 |
|  |  |  |  |  |  |  | 846 | none |  |  | 34C0000074E7 |
|  |  |  |  |  |  |  | 847 | none |  |  | 34C00000724A |
|  |  |  |  |  |  |  | 848 | none |  |  | 34C0000072D8 |
|  |  |  |  |  |  |  | 849 | none |  |  | 34C000007226 |
|  |  |  |  |  |  |  | 850 | none |  |  | 34 C 000007437 |
|  |  |  |  |  |  |  | 851 | none |  |  | 34C00000733C |
|  |  |  |  |  |  |  | 852 | none |  |  | 34C000007454 |
|  |  |  |  |  |  |  | 853 | none |  |  | 34C00000749E |
|  |  |  |  |  |  |  | 854 | none |  |  | 34C000007389 |
|  |  |  |  |  |  |  | 855 | none |  |  | 34C000007255 |
|  |  |  |  |  |  |  | 856 | none |  |  | 34C00000728E |
|  |  |  |  |  |  |  | 857 | none |  |  | 34C0000072F1 |
|  |  |  |  |  |  |  | 858 | none |  |  | 34C0000072EE |
|  |  |  |  |  |  |  | 859 | none |  |  | 34C0000073C1 |
|  |  |  |  |  |  |  | 860 | none |  |  | 34C0000073B9 |
|  |  |  |  |  |  |  | 861 | none |  |  | 34C000007490 |
|  |  |  |  |  |  |  | 862 | none |  |  | 34C000007414 |
|  |  |  |  |  |  |  | 863 | none |  |  | 34C0000072C7 |
|  |  |  |  |  |  |  | 864 | none |  |  | 34C0000073BF |
|  |  |  |  |  |  |  | 865 | none |  |  | 34C0000072F0 |
|  |  |  |  |  |  |  | 866 | none |  |  | 34C0000073C3 |
|  |  |  |  |  |  |  | 867 | none |  |  | 34C0000074F2 |

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1 Canada continued.

| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | 868 | none |  |  | 34C000007225 |
|  |  |  |  |  |  |  | 869 | none |  |  | 34C000007299 |
|  |  |  |  |  |  |  | 870 | none |  |  | 34C000007427 |
|  |  |  |  |  |  |  | 874 | none |  |  | 34C00000756C |
|  |  |  |  |  |  |  | 875 | none |  |  | 34C000007479 |
|  |  |  |  |  |  |  | 876 | none |  |  | 34C0000073CE |
|  |  |  |  |  |  |  | 877 | none |  |  | 34C0000074F7 |
|  |  |  |  |  |  |  | 878 | none |  |  | 34C000007211 |
|  |  |  |  |  |  |  | 879 | none |  |  | 34C000007366 |
|  |  |  |  |  |  |  | 880 | none |  |  | 34C0000073F8 |
|  |  |  |  |  |  |  | 881 | none |  |  | 34C0000072F6 |
|  |  |  |  |  |  |  | 882 | none |  |  | $34 \mathrm{C000007311}$ |
|  |  |  |  |  |  |  | 883 | none |  |  | 34C0000074F9 |
|  |  |  |  |  |  |  | 884 | none |  |  | 34C000007466 |
|  |  |  |  |  |  |  | 885 | none |  |  | 34C000007237 |
|  |  |  |  |  |  |  | 886 | none |  |  | 34C0000072CE |
|  |  |  |  |  |  |  | 887 | none |  |  | 34C00000746C |
|  |  |  |  |  |  |  | 888 | none |  |  | $34 \mathrm{C00000750C}$ |
|  |  |  |  |  |  |  | 889 | none |  |  | 34C000007256 |
|  |  |  |  |  |  |  | 890 | none |  |  | 34C000007329 |
|  |  |  |  |  |  |  | 891 | none |  |  | 34C00000730D |
|  |  |  |  |  |  |  | 892 | none |  |  | 34C00000752B |
|  |  |  |  |  |  |  | 893 | none |  |  | 34C0000074CB |
|  |  |  |  |  |  |  | 894 | none |  |  | 34C000007547 |
|  |  |  |  |  |  |  | 895 | none |  |  | 34C0000073DC |
|  |  |  |  |  |  |  | 896 | none |  |  | 34C000007514 |
|  |  |  |  |  |  |  | 897 | none |  |  | 34C000007570 |
|  |  |  |  |  |  |  | 898 | none |  |  | 34C00000737E |
|  |  |  |  |  |  |  | 899 | none |  |  | 34C00000742F |
|  |  |  |  |  |  |  | P2600 | none |  |  | 34C0000073D7 |
|  |  |  |  |  |  |  | P2601 | none |  |  | 34C000007213 |
|  |  |  |  |  |  |  | P2602 | none |  |  | 34C00000725B |
|  |  |  |  |  |  |  | P2603 | none |  |  | 34C0000074C5 |
|  |  |  |  |  |  |  | P2604 | none |  |  | 34C000007264 |
|  |  |  |  |  |  |  | P2605 | none |  |  | 34C000007261 |
|  |  |  |  |  |  |  | P2606 | none |  |  | 34C0000072B8 |

1 Canada continued.

| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | P2607 | none |  |  | 34C000007475 |
|  |  |  |  |  |  |  | P2608 | none |  |  | 34 C 000007352 |
|  |  |  |  |  |  |  | P2609 | none |  |  | 34 C 000007480 |
|  |  |  |  |  |  |  | P2610 | none |  |  | 34C0000072DE |
|  |  |  |  |  |  |  | P2611 | none |  |  | 34C000007390 |
|  |  |  |  |  |  |  | P2612 | none |  |  | 34C0000074A0 |
|  |  |  |  |  |  |  | P2613 | none |  |  | 34C0000072CA |
|  |  |  |  |  |  |  | P2614 | none |  |  | 34C000007563 |
|  |  |  |  |  |  |  | P2615 | none |  |  | 34C000007545 |
|  |  |  |  |  |  |  | P2616 | none |  |  | $34 \mathrm{C000007328}$ |
|  |  |  |  |  |  |  | P2617 | none |  |  | 34 C 0000074 A 1 |
|  |  |  |  |  |  |  | P2618 | none |  |  | 34C000007486 |
|  |  |  |  |  |  |  | P2619 | none |  |  | 34C000007446 |
|  |  |  |  |  |  |  | P2620 | none |  |  | $34 \mathrm{C000007518}$ |
|  |  |  |  |  |  |  | P2621 | none |  |  | 34C00000753F |
|  |  |  |  |  |  |  | P2622 | none |  |  | 34C000007481 |
|  |  |  |  |  |  |  | P2623 | none |  |  | $34 \mathrm{C000007435}$ |
|  |  |  |  |  |  |  | P2624 | none |  |  | $34 \mathrm{C000007221}$ |
|  |  |  |  |  |  |  | P2625 | none |  |  | $34 \mathrm{C000007356}$ |
|  |  |  |  |  |  |  | P2626 | none |  |  | 34 C 000007313 |
|  |  |  |  |  |  |  | P2627 | none |  |  | 34C00000745A |
|  |  |  |  |  |  |  | P2628 | none |  |  | 34C000007449 |
|  |  |  |  |  |  |  | P2629 | none |  |  | 34C0000073FF |
|  |  |  |  |  |  |  | P2630 | none |  |  | 34C0000072AF |
|  |  |  |  |  |  |  | P2631 | none |  |  | 34C0000074AC |
|  |  |  |  |  |  |  | P2632 | none |  |  | 34C00000749A |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | Kelt | W | Highlands River | I-Button Archieve | 14 | L2000 | none | May | Highland R | 34C00000761E |
|  |  |  |  |  | \& Red Atkins |  | L2001 | none |  |  | 34C000007787 |
|  |  |  |  |  |  |  | L2002 | none |  |  | 34C000007609 |
|  |  |  |  |  |  |  | L2003 | none |  |  | 34C000007606 |
|  |  |  |  |  |  |  | L2004 | none |  |  | 34C000007812 |
|  |  |  |  |  |  |  | L2005 | none |  |  | 34C0000078AD |
|  |  |  |  |  |  |  | L2006 | none |  |  | $34 \mathrm{C000007724}$ |
|  |  |  |  |  |  |  | L2007 | none |  |  | 34C00000759D |
|  |  |  |  |  |  |  | L2008 | none |  |  | $34 \mathrm{C000007707}$ |

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1 Canada continued.

| Marking <br> Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | L2009 | none |  |  | 34C0000075A7 |
|  |  |  |  |  |  |  | L2010 | none |  |  | 34C00000760B |
|  |  |  |  |  |  |  | L2011 | none |  |  | 34C000007876 |
|  |  |  |  |  |  |  | L2012 | none |  |  | 34C00000788E |
|  |  |  |  |  |  |  | L2013 | none |  |  | 34C000007625 |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | Kelts | W | Western Arm Bk | Carlin-green | 3 | N5627-N5629 | none | May-Jun | Western Arm Brook |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | Smolt | W | Conne River | STR-green | 3361 | 565-3930 | none | April- June | Conne R |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| Quebec |  | Adult | W | Escoumins | T Bar Spagetti | 14 | 03682 à 03695 | none | June 2000 | Escoumins (sea) |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| Maritimes | 2-4 | smolt | W | Tay | STR-green | 336 | 9885-9899,5650-5972 | ADC | Apr-May | Tay R |  |
|  | 2-4 | smolt | W | Tay | ADC | 329 |  | none | Apr-May | Tay R |  |
|  | 2-4 | smolt | 70W \& | Tobique | STR-green | 78 | 1311-1351 | none | May-Jun | Tobique R - Odell |  |
|  |  |  | 8H |  |  |  | 1763-1800 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | 2-4 | smolt | 6W,16H | Tobique | STR-green | 22 | 2066-2087 | none | May-Jun | Tobique R - Headpond |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0+ to 3+ | parr | 279W\& | Tobique | STR-green | 334 | 5550-5798,9001-9205 | none | Oct-Nov | Tobique R - Nictau |  |
|  |  |  | 55H |  |  |  | 18561-18605,799 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0+ to 3+ | parr | 350W\& | Tobique | STR-green | 420 | 5973-6099,6350-6449 | none | Oct-Nov | Tobique R - Gulquac |  |
|  |  |  | 70 H |  |  |  | 15100-15145 |  |  |  |  |
|  |  |  |  |  |  |  | 18700-18749 |  |  |  |  |
|  |  |  |  |  |  |  | 18850-19949 |  |  |  |  |
|  | 0+ to 3+ | parr | 463W\& | Tobique | STR-green | 556 | 2088-2100,6100-6349 | none | Oct-Nov | Tobique R - Three Brooks |  |
|  |  |  | 93H |  |  |  | 6450-6499 |  |  |  |  |
|  |  |  |  |  |  |  | 15000-15092 |  |  |  |  |
|  |  |  |  |  |  |  | 18501-18549 |  |  |  |  |
|  |  |  |  |  |  |  | 18800-18849 |  |  |  |  |
|  |  |  |  |  |  |  | 18950-18999 |  |  |  |  |
|  | 1SW | adult | 4W \& 1H | Hammond | Carlin-yellow | 5 | 4765,67,90-92 | none | Aug | Hammond R | Estuary |
|  | 1SW/MSW | adult | 13W\&3H | Hammond | Carlin-green | 16 | 47106-47199 | none | Nov | Hammond R | Broodstock |
|  | 1SW/MSW | adult | W | Nashwaak | Carlin-white | 18 | 18590-18614 | none | Nov | Nashwaak | Broodstock |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1+ | smolt | H | Annapolis | ADC | 2396 |  | none | Apr-May | Annapolis R |  |
|  | 1+ | smolt | H | Tusket | ADC | 10840 |  | none | Apr-May | Bear R |  |

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1 Canada continued.

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| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $1+$ | smolt | H | LaHave | ADC | 9137 |  | none | Apr-May | Clyde R |  |
|  | 1+ | parr | H | East R Sh Hbr | ADC | 1350 |  | none | Apr-May | East R Sh Hbr |  |
|  | 0+ | parr | H | East R Sh Hbr | ADC | 20000 |  | none | Oct | East R Sh Hbr |  |
|  | 1+ | smolt | H | East R Sh Hbr | ADC | 1650 |  | none | Apr-May | East R Sh Hbr |  |
|  | 2+ | smolt | H | East R Sh Hbr | STR-red,blue | 6768 |  | ADC | Apr-May | East R Sh Hbr |  |
|  | 0+ | parr | H | Gaspereau, Kings | ADC | 21880 |  | none | Oct | Gaspereau, Kings Co |  |
|  | $1+$ | smolt | H | Gaspereau, Kings | ADC | 15874 |  | none | Apr-May | Gaspereau, Kings Co |  |
|  | 0+ | parr | H | Gold | ADC | 8748 |  | none | Oct | Gold R |  |
|  | 2+ | smolt | H | Gold | ADC | 11656 |  | none | Apr-May | Gold R |  |
|  | 1+ | smolt | H | LaHave | ADC | 4494 |  | none | Apr-May | Jordan R |  |
|  | 0+ | parr | H | LaHave | ADC | 7878 |  | none | Apr-May | LaHave R |  |
|  | 0+ | parr | H | LaHave | ADC | 72740 |  | none | Oct | LaHave R |  |
|  | $1+$ | smolt | H | LaHave | ADC | 30109 |  | none | Apr-May | LaHave R |  |
|  | $1+$ | smolt | H | LaHave | STR-clear,blue, | 19999 |  | ADC | Apr-May | LaHave R |  |
|  |  |  |  |  | red, green |  |  |  |  |  |  |
|  | $1+$ | parr | H | Liscomb | ADC | 7375 |  | none | Apr-May | Liscomb R |  |
|  | $1+$ | smolt | H | Liscomb | ADC | 10020 |  | none | Apr-May | Liscomb R |  |
|  | 0+ | parr | H | Liscomb | ADC | 19496 |  | none | Oct | Liscomb R |  |
|  | 0+ | parr | H | Margaree | ADC | 31936 |  | none | Oct | Margaree R |  |
|  | 0+ | parr | H | Medway | ADC | 13216 |  | none | Oct | Medway R |  |
|  | $1+$ | smolt | H | Medway | ADC | 50944 |  | none | Apr-May | Medway R |  |
|  | 1+ | smolt | H | LaHave | ADC | 10048 |  | none | Apr-May | Mersey R |  |
|  | 0+ | parr | H | Tusket | ADC | 14360 |  | none | Oct | Meteghan R |  |
|  | 0+ | parr | H | LaHave | ADC | 13600 |  | none | Oct | Mushamush R |  |
|  | 1+ | smolt | H | LaHave | ADC | 7998 |  | none | Apr-May | Mushamush R |  |
|  | 1+ | parr | H | Musquodoboit | ADC | 4837 |  | none | Apr-May | Musquodoboit R |  |
|  | 0+ | parr | H | Musquodoboit | ADC | 17440 |  | none | Oct | Musquodoboit R |  |
|  | $1+$ | smolt | H | Musquodoboit | ADC | 14404 |  | none | Apr-May | Musquodoboit R |  |
|  | 0+ | parr | H | LaHave | ADC | 14000 |  | none | Oct | Petite Riviere |  |
|  | $1+$ | smolt | H | LaHave | ADC | 8085 |  | none | Apr-May | Petite Riviere |  |
|  | 0+ | parr | H | Sackville | ADC | 25416 |  | none | Oct | Sackville R |  |
|  | $1+$ | smolt | H | Sackville | ADC | 23964 |  | none | Apr-May | Sackville R |  |
|  | 0+ | parr | H | Salmon R Digby | ADC | 30520 |  | none | Oct | Salmon R Digby |  |
|  | 1+ | smolt | H | Salmon R Digby | ADC | 25596 |  | none | Apr-May | Salmon R Digby |  |
|  | 0+ | parr | H | Tusket | ADC | 13622 |  | none | Oct | Tusket R |  |
|  | 1+ | smolt | H | Tusket | ADC | 49661 |  | none | Apr-May | Tusket R |  |

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1 Canada continued.

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| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $1+$ | smolt | H | Tusket | STR-clear,blue | 8000 |  | ADC | Apr-May | Tusket R |  |
|  | 1+ | smolt | H | Saint John | Carlin-blue | 10000 | V48000-49999 | ADC | Apr-May | Saint John R |  |
|  |  |  |  |  |  |  | V37500-39999 | ADC |  |  |  |
|  |  |  |  |  |  |  | V41000-42999 | ADC |  |  |  |
|  |  |  |  |  |  |  | V44500-44999 | ADC |  |  |  |
|  |  |  |  |  |  |  | EE59000-59999 | ADC |  |  |  |
|  |  |  |  |  |  |  | F20000-21999 | ADC |  |  |  |
|  | $1+$ | smolt | H | Saint John | ADC | 321136 |  | none | Apr-May | Saint John R |  |
|  | 0+ | parr | H | Saint John | ADC | 387042 |  | none | Oct | Saint John R |  |
|  |  | Adult | 165 W | Saint John | Carlin-yellow | 165 | 4305-4307 | none | Jun-Aug | Saint John R |  |
|  |  |  |  |  |  |  | 4313-4315 |  |  |  |  |
|  |  |  |  |  |  |  | 4509-4512 |  |  |  |  |
|  |  |  |  |  |  |  | 4599-4608 |  |  |  |  |
|  |  |  |  |  |  |  | 4610-4634 |  |  |  |  |
|  |  |  |  |  |  |  | 4636-4699 |  |  |  |  |
|  |  |  |  |  |  |  | 4848-4852 |  |  |  |  |
|  |  |  |  |  |  |  | 4856-4858 \& |  |  |  |  |
|  |  |  |  |  |  |  | 4319,4321,4323, |  |  |  |  |
|  |  |  |  |  |  |  | 4356,4380,4381, |  |  |  |  |
|  |  |  |  |  |  |  | 4384,4386,4387, |  |  |  |  |
|  |  |  |  |  |  |  | 4389,4500,4506, |  |  |  |  |
|  |  |  |  |  |  |  | 4526,4527,4535, |  |  |  |  |
|  |  |  |  |  |  |  | 4547,4550,4552, |  |  |  |  |
|  |  |  |  |  |  |  | 4555,4574,4582, |  |  |  |  |
|  |  |  |  |  |  |  | 4588,4591,4596, |  |  |  |  |
|  |  |  |  |  |  |  | 4598,4805,4822, |  |  |  |  |
|  |  |  |  |  |  |  | 4825,4828,4829, |  |  |  |  |
|  |  |  |  |  |  |  | 4832,4835,4836, |  |  |  |  |
|  |  |  |  |  |  |  | 4839,4840,4842, |  |  |  |  |
|  |  |  |  |  |  |  | 4844,4845,4846, |  |  |  |  |
|  |  |  |  |  |  |  | 4854,4867,4871, |  |  |  |  |
|  |  |  |  |  |  |  | 4891,4894,4898, |  |  |  |  |
|  |  |  |  |  |  |  | 489,914,635.00 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | Adult | 10H\& | Saint John | Carlin-blue | 40 | 2700-2739 | none | Nov | Saint John R | Broodstock |
|  |  |  | 30W |  |  |  |  |  |  |  |  |

1 Canada continued.

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| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Adult | 20 H | Saint John | Carlin-blue | 20 | 2850-2868\&2899 | none | July | Saint John R |  |
|  |  | Adult | 19H | Saint John | Carlin-blue | 19 | 1280-1299 | none | Oct | Saint John R |  |
|  |  | Adult | 13H\& | Saint John | Carlin-white | 68 | 18700-18752 | none | Nov | Saint John R | Broodstock |
|  |  |  | 55W |  |  |  | 18754-18762 |  |  |  |  |
|  |  |  |  |  |  |  | 18547,18764,19372, |  |  |  |  |
|  |  |  |  |  |  |  | 19400,19468-69 |  |  |  |  |
|  |  | Adult | 45 H | Saint John | Carlin-white | 45 | 18642-18645 | none |  |  |  |
|  |  |  |  |  |  |  | 18647-18649 |  |  |  |  |
|  |  |  |  |  |  |  | 18651-18652 |  |  |  |  |
|  |  |  |  |  |  |  | 18662-18663 |  |  |  |  |
|  |  |  |  |  |  |  | 18671-18683 |  |  |  |  |
|  |  |  |  |  |  |  | 18688-18691 \& |  |  |  |  |
|  |  |  |  |  |  |  | 18627,18634,18636, |  |  |  |  |
|  |  |  |  |  |  |  | 18638,18640,18654, |  |  |  |  |
|  |  |  |  |  |  |  | 18655,18660,18665, |  |  |  |  |
|  |  |  |  |  |  |  | 18666,18668,18669, |  |  |  |  |
|  |  |  |  |  |  |  | 18686,18693,18694, |  |  |  |  |
|  |  |  |  |  |  |  | 1,869,618,697 |  |  |  |  |
|  |  | Adult | 19H\& | Saint John | Carlin-red | 55 | 4008-4061,4099 | none | Nov | Saint John R | Broodstock |
|  |  |  | 40W |  |  |  |  |  |  |  |  |
|  |  | Adult | 39H\& | Saint John | Carlin-orange | 97 | 3000-3002 | none | Jun-Oct | Saint John R |  |
|  |  |  | 58W |  |  |  | 3007-3010 |  |  |  |  |
|  |  |  |  |  |  |  | 3088-3091 |  |  |  |  |
|  |  |  |  |  |  |  | 3093-3103 |  |  |  |  |
|  |  |  |  |  |  |  | 3105-3119 |  |  |  |  |
|  |  |  |  |  |  |  | 3121-3123 |  |  |  |  |
|  |  |  |  |  |  |  | 3135-3137 |  |  |  |  |
|  |  |  |  |  |  |  | 3139-3146 |  |  |  |  |
|  |  |  |  |  |  |  | 3148-3165 |  |  |  |  |
|  |  |  |  |  |  |  | 3169-3175 |  |  |  |  |
|  |  |  |  |  |  |  | 3177-3179 |  |  |  |  |
|  |  |  |  |  |  |  | 3190-3194 |  |  |  |  |
|  |  |  |  |  |  |  | 3197-3198 |  |  |  |  |
|  |  |  |  |  |  |  | 3005,3125,3127, |  |  |  |  |
|  |  |  |  |  |  |  | 3113,3167,3181, |  |  |  |  |
|  |  |  |  |  |  |  | 3183,3185,3186, |  |  |  |  |

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1 Canada continued.

| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary <br> Tag or Mark | $\begin{gathered} \text { Release } \\ \text { Date } \end{gathered}$ | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | 3188 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| St. Croix | 1SW/MSW | Adult | H | St. Croix | Carlin-blue | 10 | A39833, 34, 37,45, | ADC | Nov | St. Croix R | broodstock |
| Waterway |  |  |  |  |  |  | 77, 78, 79,83, 85, |  |  |  |  |
| Commission |  |  |  |  |  |  | and 86 |  |  |  |  |
|  |  | Adult | H | St. Croix | Carlin-blue | 48 | E97900,01,05,11, | ADC | Oct | St. Croix R | captively-reared |
|  |  |  |  |  |  |  | 12,13,17,19,23,24, |  |  |  | adults, released |
|  |  |  |  |  |  |  | 26,27,30,32,37,40, |  |  |  | as spawners |
|  |  |  |  |  |  |  | 41,42,43,44,46,50, |  |  |  |  |
|  |  |  |  |  |  |  | 54,57,58,60,61,62, |  |  |  |  |
|  |  |  |  |  |  |  | 64,68,69,71,73,74, |  |  |  |  |
|  |  |  |  |  |  |  | 76,77,80,81,86,90, |  |  |  |  |
|  |  |  |  |  |  |  | 91,92,95,96,98,99 |  |  |  |  |
|  |  |  |  |  |  |  | and 1 unknown \# |  |  |  |  |

## ICES COMPILATION OF MICROTAG, FIN CLIP, AND EXTERNAL TAG RELEASES

Marking Season: 2000

| Origin | Primary Tag or M ark |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
|  | Microtag | External <br> mark | Adipose <br> clip | Total |
| Hatchery juvenile | 72,900 | 0 | 0 |  |
| Wild juvenile | 0 | 0 | 0 | 0 |
| Adult | 0 | 0 | 0 | 0 |
| Total | 72,900 | 0 | 0 | 72,900 |


| Marking <br> Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or <br> Mark** | Number <br> Marked | Code or Serial | Secondary Tag <br> or Mark* | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DIFR | 1 | smolt | H |  | CWT | 72900 |  | ADC |  | Jutland |  |

ICES COMPILATION OF MICROTAG, FIN CLIP, AND EXTERNAL TAG RELEASES
Marking Season: 2000
Country: Iceland

| Origin | Primary Tag or Mark |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: |
|  | Microtag | External <br> mark | Adipose <br> clip | Total |
| Hatchery juvenile | 127,162 | 0 | 0 |  |
| Wild juvenile | 2,516 | 0 | 0 | 2,516 |
| Adult | 0 | 563 | 0 | 563 |
| Total | 129,678 | 563 | 0 | 130,241 |


| Marking Agency* | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark** | Number Marked | Code or Serial | Secondary Tag or Mark** | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| IFF | 1+ | smolt | H | Laxá í Adaldal | DCWT | 3112 | 016463ser | ADC | June/July | Breiơdalsá |  |
| IFF | 1+ | smolt | H | Hrútafjaròará | DCWT | 1990 | 016463ser | ADC | June/July | Hrútafjarðará |  |
| IFF | 1+ | smolt | H | Laxá í Að̌aldal | DCWT | 3007 | 016463ser | ADC | June/July | Laxá í Adaldal |  |
| IFF | 1+ | smolt | H | Miơfjarð̃ará | DCWT | 2612 | 016463ser | ADC | June/July | Miơfjarðará |  |
| IFF | 1+ | smolt | H | Koll | DCWT | 1002 | 016563ser | ADC | June/July | Eystri Rangá |  |
| IFF | 1+ | smolt | H | Laxá í Leirársveit | DCWT | 7535 | 016563ser | ADC | June/July | Laxá í Leirársveit |  |
| IFF | 1+ | smolt | H | Norơurá | DCWT | 3009 | 016563ser | ADC | June/July | Norơurá |  |
| IFF | 1+ | smolt | H | Elliðaár | DCWT | 1069 | 016663ser | ADC | June/July | Elliõaár |  |
| IFF | $1+$ | smolt | H |  | DCWT | 990 | 016663ser | ADC | June/July | Eystri Rangá |  |
| IFF | 1+ | smolt | H | Sogiơ | DCWT | 4034 | 016663ser | ADC | June/July | Sogió |  |
| IFF | $1+$ | smolt | H | Dalir | DCWT | 1000 | 016663ser | ADC | June/July | Ytri Rangá |  |
| IFF | 1+ | smolt | H | Koll | DCWT | 3002 | 016663ser | ADC | June/July | Ytri Rangá |  |
| IFF | $1+$ | smolt | H | Norơruá | DCWT | 1035 | 016663ser | ADC | June/July | Ytri Rangá |  |
| IFF | 1+ | smolt | H | Koll | DCWT | 9037 | 016763ser | ADC | June/July | Ytri Rangá |  |
| IFF | 1+ | smolt | H | Rangár | DCWT | 3021 | 016763ser | ADC | June/July | Ytri Rangá |  |
| IFF | 1+ | smolt | H | Breiơdalsá | DCWT | 1010 | 016863ser | ADC | June/July | Breiôdalsá |  |
| IFF | $1+$ | smolt | H | Koll | DCWT | 4012 | 016863ser | ADC | June/July | Eystri Rangá |  |
| IFF | 1+ | smolt | H | Rangár | DCWT | 3022 | 016863ser | ADC | June/July | Eystri Rangá |  |
| IFF | 1+ | smolt | H | Koll | DCWT | 1005 | 016863ser | ADC | June/July | Ytri Rangá |  |

3 Iceland continued.

| Marking Agency* | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark** | Number Marked | Code or Serial | Secondary Tag or Mark** | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| IFF | 1+ | smolt | H | Koll | DCWT | 2023 | 016863ser | ADC | June/July | Pverá |  |
| IFF | 1+ | smolt | H | Koll | DCWT | 2004 | 016963ser | ADC | June/July | Eystri Rangá |  |
| IFF |  | smolt | W | Vesturdalsá | DCWT | 1313 | 016963ser | ADC | June/July | Vesturdalsá |  |
| IFF | 1+ | smolt | H | Koll | DCWT | 1013 | 016963ser | ADC | June/July | Ytri Rangá |  |
| IFF |  | smolt | W | Austurá | DCWT | 401 | 017063ser | ADC | June/July | Austurá |  |
| IFF |  | smolt | W | Elliòaár | DCWT | 802 | 017063ser | ADC | June/July | Elliõaár |  |
| IFF | 0+ | smolt | H | Pjórsá | CWT | 340 | 064263ser | ADC | June/July | Pjórsá |  |
| IFF | 0+ | smolt | H | Laxá í Adaldal | CWT | 1010 | 064663ser | ADC | June/July | Laxá í Aďaldal |  |
| IFF | 0+ | smolt | H | Pjórsá | CWT | 1660 | 064763ser | ADC | June/July | Pjórsá |  |
| IFF | 0+ | smolt | H | Elliozaár | CWT | 5163 | 064963ser | ADC | June/July | Ellið̃avatn |  |
| IFF | 0+ | smolt | H | Pjórsá | CWT | 6998 | 065263ser | ADC | June/July | Pjórsá |  |
| IFF | 0+ | smolt | H | Laxá í Adaldal | CWT | 6990 | 065363ser | ADC | June/July | Laxá í Aďaldal |  |
| IFF | 0+ | smolt | H | Pjórsá | CWT | 1444 | 065363ser | ADC | June/July | Pjórsá |  |
| IFF | $1+$ | smolt | H | Elliõaár | CWT | 4527 | 065463ser | ADC | June/July | Elliõaár |  |
| IFF | 0+ | smolt | H | Elliðaár | CWT | 6849 | 065463ser | ADC | June/July | Elliðavatn |  |
| IFF | $1+$ | smolt | H | Elliozaár | CWT | 5166 | 065563ser | ADC | June/July | Elliõaár |  |
| IFF | 0+ | smolt | H | Sogiô | CWT | 6015 | 065563ser | ADC | June/July | Sogiơ |  |
| IFF | $1+$ | smolt | H | Elliozaár | CWT | 337 | 065663ser | ADC | June/July | Elliòaár |  |
| IFF | 1+ | smolt | H | Fnjóská | CWT | 10013 | 065663ser | ADC | June/July | Fnjóská |  |
| IFF | $1+$ | smolt | H | Elliòaár | CWT | 11106 | 065763ser | ADC | June/July | Elliòaár |  |
| IFF |  | adult |  |  | T_Bar_green | 396 | IS71000-IS71499 | none | July-Aug | Norolingafljót | Put \& Take |
| IFF |  | adult |  |  | T_Bar_green | 44 | IS71000-IS71499 | none | July-Aug | Hofsá | Catch \& Rel. |
| IFF |  | adult |  | Laxá í Adaladal | T_Bar_white | 20 | IS52775-IS52794 | none | Nov. 1999 | Laxá í Adaladal | Kelt |
| IFF |  | adult |  | Laxá í Adaladal | T_Bar_white | 1 | IS52819 | none | November | Laxá í Adaladal | Kelt |
| IFF |  | adult |  | Laxá í Adaladal | T_Bar_white | 3 | IS52822-IS52824 | none | November | Laxá í Adaladal | Kelt |
| IFF |  | adult |  | Laxá í Adaladal | T_Bar_green | 2 | IS58548-IS58549 | none | November | Laxá í Adaladal | Kelt |
| IFF |  | adult |  | Laxá í Adaladal | T_Bar_green | 1 | IS61325 | none | November | Laxá í Adaladal | Kelt |
| IFF |  | adult |  | Laxá í Adaladal | T_Bar_green | 12 | IS61358-IS61369 | none | November | Laxá í Adaladal | Kelt |
| IFF |  | adult |  | Laxá í Adaladal | T_Bar_white | 10 | 01416-01425 | none | November | Laxá í Adaladal | Kelt |
| IFF |  | adult |  | Lake Hóp | T_Bar_white | 1 | 2564 | none | June | Lake Hóp | Kelt |
| IFF |  | adult |  | Lake Hóp | T_Bar_white | 1 | 2598 | none | June | Lake Hóp | Kelt |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_white | 1 | IS54797 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_white | 1 | IS54787 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_white | 1 | IS54751 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_white | 1 | IS57847 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Sog | T_Bar_white | 1 | IS54780 | none | December | Ölfusá | Kelt |


| 3 Iceland continued. |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Marking <br> Agency* | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark** | Number Marked | Code or Serial | Secondary Tag or Mark** | Release Date | Release Location | Comment |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_white | 1 | IS54820 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_white | 1 | IS54848 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Sog | T_Bar_white | 1 | IS54782 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_white | 1 | IS54821 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_white | 1 | IS2883 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_white | 1 | IS54819 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_white | 1 | IS54788 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Sog | T_Bar_white | 1 | IS54785 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Sog | T_Bar_white | 1 | IS54756 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Sog | T_Bar_white | 1 | IS54783 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_white | 1 | IS54849 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_white | 1 | IS54789 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_white | 1 | IS2882 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Stóra-Laxá | T_Bar_green | 1 | IS69699 | none | December | Ölfusá | Kelt |
| IFF |  | adult |  | Ölfusá | T_Bar_white | 1 | IS2820 | none | July | Ölfusá | Kelt |
| IFF |  | adult |  | Ölfusá | T_Bar_white | 1 | IS2822 | none | July | Ölfusá | Kelt |
| IFF |  | adult |  | Ölfusá | T_Bar_white | 1 | IS54803 | none | July | Ölfusá | Kelt |
| IFF |  | adult |  |  | T_Bar_blue | 50 | Grimsa 2000-3000 | none | July-Aug | Grímsá | Catch \& Rel. |

* Institute of Freshwater Fisheries
** DCWT =Decimal coded wire tag, CWT= coded wire tag, ADC=adipose fin clip


## ICES COMPILATION OF MICROTAG, FIN CLIP, AND EXTERNAL TAG RELEASES

Marking Season: 2000

|  | Primary Tag or Mark |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
| Orig in | Microtag | External <br> mark | Adipose <br> clip | Total |
| Hatchery juvenile | 289,029 | 0 | 0 |  |
| Wild juvenile | 939 | 0 | 0 | 939 |
| Adult | 0 | 0 | 0 | 0 |
| Total | 289,968 | 0 | 0 | 289,968 |


| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Tom McDermott | 1+ | Parr | H | Shannon | CWT | 10127 | 47/03/12 | Adipose | 04/26/2000 | Shannon | MSW/Gr. Study |
| Marine Institute | 1+ | Parr | H | Shannon | CWT | 8060 | 47/19/62A | Adipose | 04/26/2000 | Shannon | MSW/Gr. Study |
| Ireland | 1+ | Parr | H | Shannon | CWT | 2037 | 47/16/33D | Adipose | 04/25/2000 | Shannon | MSW/Gr. Study |
| Migration 2000 | 1+ | Parr | H | Shannon | CWT | 10095 | 47/03/13 | Adipose | 04/25/2000 | Shannon | MSW/Gr. Study |
|  | 1+ | Parr | H | Shannon | CWT | 10139 | 47/03/14 | Adipose | 04/25/2000 | Shannon | MSW/Gr. Study |
|  | $1+$ | Parr | H | Shannon | CWT | 10997 | 47/03/15 | Adipose | 04/27/2000 | Shannon | MSW |
|  | 1+ | Parr | H | Shannon | CWT | 11028 | 47/04/10 | Adipose | 04/27/2000 | Shannon | MSW |
|  | 1+ | Parr | H | Shannon | CWT | 11004 | 47/04/09 | Adipose | 04/27/2000 | Shannon | MSW |
|  | $1+$ | Parr | H | Shannon | CWT | 6145 | 47/04/08 | Adipose | 04/27/2000 | Shannon | MSW |
|  | $1+$ | Parr | H | Shannon | CWT | 7274 | 47/19/56 | Adipose | 04/26/2000 | Liffey | transfer - mix |
|  | $1+$ | Parr | H | Cong | CWT | 11,022 | 47/02/08 | Adipose | 03/27/2000 | Corrib | Cong river |
|  | 1+ | Parr | H | Ballynahinch | CWT | 4019 | 47/19/54A | Adipose | 04/14/2000 | Owenmore | transfer |
|  | $1+$ | Parr | H | Ballynahinch | CWT | 5680 | 47/19/55A | Adipose | 04/14/2000 | Owenmore | transfer |
|  | 1+ | Parr | H | Bunowen | CWT | 11301 | 47/02/06 | Adipose | 03/28/2000 | Bunowen | transfer |
|  | $1+$ | Parr | H | Bunowen | CWT | 991 | 47/19/35A | Adipose | 03/28/2000 | Bunowen | transfer |
|  | $1+$ | Parr | H | Lee | CWT | 10827 | 47/04/01 | Adipose | 03/30/2000 | Inniscarra | indigenous |
|  | $1+$ | Parr | H | Burrishoole | CWT | 10997 | 47/04/02 | Adipose | 05/01/2000 | Bundorragha | non-indigenous |
|  | 1+ | Parr | H | Burrishoole | CWT | 9237 | 47/19/61A | Adipose | 05/01/2000 | Bundorragha | non-indigenous |
|  | $1+$ | Parr | H | Burrishoole | CWT | 2772 | 47/18/21B | Adipose | 05/01/2000 | Bundorragha | non-indigenous |

Ireland continued.

| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1+ | Parr | H | Delphi | CWT | 11004 | 47/04/07 | Adipose | 05/01/2000 | Bundorragha | indigenous |
|  | $1+$ | Parr | H | Delphi | CWT | 3601 | 47/20/02A | Adipose | 05/01/2000 | Bundorragha | indigenous |
|  | 1+ | Parr | H | Delphi | CWT | 11018 | 47/04/03 | Adipose | 05/01/2000 | Bundorragha | indigenous |
|  | $1+$ | Parr | H | Delphi | CWT | 5620 | 47/19/37A | Adipose | 05/01/2000 | Bundorragha | indigenous |
|  | 1+ | Parr | H | Delphi | CWT | 10724 | 47/04/04 | Adipose | 05/01/2000 | Bundorragha | indigenous |
|  | $1+$ | Parr | H | Delphi | CWT-seq. | 6412 | 47/02/01B | Adipose | 05/01/2000 | Bundorragha | indigenous |
|  | 1+ | Parr | H | Shannon | CWT/ X | 4030 | 47/19/39A | Adipose | 04/28/2000 | Burrishoole | MSW/Gr. Study |
|  | $1+$ | Parr | H | Shannon | CWT/ S | 2049 | 47/18/26B | Adipose | 04/28/2000 | Burrishoole | MSW/Gr. Study |
|  | $1+$ | Parr | H | Shannon | CWT/ S | 2007 | 47/18/25B | Adipose | 04/28/2000 | Burrishoole | MSW/Gr. Study |
|  | $1+$ | Parr | H | Shannon | CWT/ S | 266 | 47/19/53A | Adipose | 04/28/2000 | Burrishoole | MSW/Gr. Study |
|  | $1+$ | Parr | H | Shannon | CWT/ T | 1437 | 47/18/63A | Adipose | 04/28/2000 | Burrishoole | MSW/Gr. Study |
|  | $1+$ | Parr | H | Shannon | CWT/ T | 720 | 47/03/41K | Adipose | 04/28/2000 | Burrishoole | MSW/Gr. Study |
|  | $1+$ | Parr | H | Shannon | CWT/ T | 1337 | 47/17/63B | Adipose | 04/28/2000 | Burrishoole | MSW/Gr. Study |
|  | 1+ | Parr | H | Shannon | CWT/ H | 2063 | 47/17/23C | Adipose | 04/28/2000 | Burrishoole | MSW/Gr. Study |
|  | $1+$ | Parr | H | Shannon | CWT/ H | 1750 | 47/17/54B | Adipose | 04/28/2000 | Burrishoole | MSW/Gr. Study |
|  | 1+ | Parr | H | Shannon | CWT/ H | 274 | 47/17/33C | Adipose | 04/28/2000 | Burrishoole | MSW/Gr. Study |
|  | $1+$ | Parr | H | Shannon | CWT/ O | 8269 | 47/01/68 | Adipose | 04/28/2000 | Burrishoole | MSW/Gr.-all female |
|  | 1+ | Parr | H | Burrishoole | CWT | 7989 | 47/01/69 | Adipose | 05/02/2000 | Burrishoole | indigenous-control |
|  | $1+$ | Parr | H | Burrishoole | CWT | 8000 | 47/01/70 | Adipose | 05/02/2000 | Burrishoole | indigenous-Abbey |
|  | $1+$ | Parr | H | Burrishoole | CWT | 8984 | 47/01/71 | Adipose | 04/28/2000 | Burrishoole | indigenous |
|  | $1+$ | Parr | H | Erne | CWT | 10506 | 47/01/84 | Adipose | 05/04/2000 | Erne | Tailrace |
|  | 1+ | Parr | H | Erne | CWT | 1729 | 47/18/29B |  | 03/28/2000 | Erne | Termon |
|  | $1+$ | Parr | H | Erne | CWT | 688 | 47/18/23B |  | 03/28/2000 | Erne | Termon |
|  | $1+$ | Parr | H | Erne | CWT | 1688 | 47/19/28A | Adipose | 03/28/2000 | Erne | Waterfoot |
|  | $1+$ | Parr | H | Erne | CWT | 827 | 47/18/24B | Adipose | 03/28/2000 | Erne | Waterfoot |
|  | $1+$ | Parr | H | Erne | CWT | 9189 | 47/01/85 | Adipose | 05/04/2000 | Erne | Tailrace |
|  | 1+ | Parr | H | Screebe | CWT | 11389 | 47/01/86 | Adipose | 04/04/2000 | Screebe | Waterfall |
|  | $1+$ | Parr | H | Coomhola | CWT | 816 | 47/17/60B | Adipose | 04/01/2000 | Coomhola |  |
|  | 1+ | Parr | H | Coomhola | CWT | 475 | 47/17/58B | Adipose | 04/01/2000 | Coomhola |  |
|  | 1+ | Parr | H | Ilen | CWT | 782 | 47/18/16B | Adipose | 04/01/2000 | Ilen |  |
|  | 1+/3+ | Smolt | W | Corrib | CWT | 573 | 47/19/38 | Adipose | 05/06/2000 | Galway | Trap |
|  | 1+/3+ | Smolt | W | Corrib | CWT | 366 | 47/17/37B | Adipose | 05/11/2000 | Galway | Trap |

5 Norway
ICES COMPILATION OF MICROTAG, FIN CLIP, AND EXTERNAL TAG RELEASES
Marking Season: 2000
Country: Norway

|  | Primary Tag or Mark |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
| Origin | Microtag | External <br> mark | Adipose <br> clip | Total |
| Hatchery juvenile | 0 | 85,692 | 0 |  |
| Wild juvenile | 0 | 5,436 | 0 | 5,436 |
| Adult | 0 | 631 | 0 | 631 |
| Total | 0 | 91,759 | 0 | 91,759 |


| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| NINA | 2 | smolt | H | Imsa | Carlin | 1993 | NG-20000-21999 |  | 02 May | Imsa |  |
| NINA | 2 | smolt | H | Figgo | Carlin | 1499 | NG-22000-23499 |  | 02 May | Imsa |  |
| NINA | 2 | smolt | H | Lone | Carlin | 1491 | NG-23500-24999 |  | 02 May | Imsa |  |
| NINA | 1 | smolt | H | Imsa | Carlin | 1989 | NG-25000-26999 |  | 02 May | Imsa |  |
| NINA | 1 | smolt | H | Figgo | Carlin | 1493 | NG-27000-28499 |  | 02 May | Imsa |  |
| NINA | 1 | smolt | H | Lone | Carlin | 1482 | NG-28500-29999 |  | 02 May | Imsa |  |
| NINA |  | smolt | W | Imsa | Carlin | 586 | NC-12229-12815 |  |  | Imsa/fella |  |
| NINA |  | smolt | W | Figgo | Carlin | 600 | NE-3400-3999 |  | 26 Apr | Figgjo |  |
| NINA |  | smolt | W | Figgjo | Carlin | 397 | NE-4600-4900 |  | 26 Apr | Figgjo |  |
| NINA | 1 | smolt | H | Figgo | Carlin | 1990 | NG-34000-35999 |  | 03 May | Figgjo |  |
| NINA | 2 | smolt | H | Imsa | Carlin | 1994 | NG-30000-31999 |  | 16 May | Dirdal |  |
| NINA | 2 | smolt | H | Imsa | Carlin | 1982 | NG-32000-33999 |  | 15 May | Frafjord |  |
| NINA | 1 | smolt | H | Suldal | Carlin | 1068 | NG-36000-37199 |  | 08 May | Imsa |  |
| NINA | 1 | smolt | H | Imsa | Carlin | 929 | NG-37200-38399 |  | 08 May | Imsa |  |
| NINA | 1 | smolt | H | Suldal | Carlin | 685 | NG-38400-39599 |  | 08 May | Imsa |  |
| NINA | 1 | smolt | H | Imsa | Carlin | 851 | NG-39600-40799 |  | 08 May | Imsa |  |
| NINA | 1 | smolt | H | Vikja | Carlin | 29999 | NF-42000-44999 |  | 04 May | Vikja |  |
| NINA | 2 | smolt | H | Eira | Carlin | 1995 | NF-45000-46999 |  | 16 May | Eira |  |
| NINA | 2 | smolt | H | Eira | Carlin | 998 | NF-62000-62999 |  | 16 May | Eira |  |


| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary <br> Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| NINA | 2 | smolt | H | Eira | Carlin | 2984 | NF-63000-65999 |  | 16 May | Eira |  |
| NINA |  | smolt | W | Audna | Carlin | 877 | NF-81000-81899 |  | 27 Mar | Audna |  |
| NINA | 1 | smolt | H | Audna | Carlin | 1984 | Nf-88000-89999 |  | 11 May | Audna |  |
| NINA |  | smolt | W | Bjerkreim | Carlin | 1000 | NF-82000-82999 |  | 04 Apr | Bjrekreim |  |
| NINA |  | smolt | W | Bjerkreim | Carlin | 976 | NF-83000--83999 |  | 03 Apr | Bjrekreim |  |
| NINA | 1 | smolt | H | Mandal | Carlin | 2957 | NF-53000-55999 |  | 11 May | Mandal |  |
| NINA | 1 | smolt | H | Mandal | Carlin | 1732 | NF-56000-57745 |  | 11 May | Mandal |  |
| NINA | 2 | smolt | H | Mandal | Carlin | 264 | NF-57746-57999 |  | 11 May | Mandal |  |
| NINA | 1 | smolt | H | Mandal | Carlin | 430 | NF-58000-58429 |  | 11 May | Mandal |  |
| NINA | 2 | smolt | H | Mandal | Carlin | 567 | NF-58430-58999 |  | 11 May | Mandal |  |
| NINA | 2 | smolt | H | Mandal | Carlin | 2893 | NF-59000-61899 |  | 12 May | Mandal |  |
| NINA | 1 | smolt | H | Drammen | Carlin | 999 | NF-84000-84999 |  | 10 May | Drammen |  |
| NINA | 1 | smolt | H | Drammen | Carlin | 996 | NF-85000-85999 |  | 10 May | Drammen |  |
| NINA | 1 | smolt | H | Drammen | Carlin | 1000 | NF-86000-86999 |  | 10 May | Drammen |  |
| NINA | 1 | smolt | H | Drammen | Carlin | 998 | NF-87000-87999 |  | 10 May | Drammen |  |
| NINA |  | smolt | W | Vosso | Carlin | 1000 | NF-91000-91999 |  | 11 Apr | Vosso |  |
| NINA | 2 | smolt | H | Alta | Carlin | 343 | NE-56653-56999 |  | 23 Jun | Halselv |  |
| NINA | 2 | smolt | H | Alta | Carlin | 106 | NF-66000-66105 |  | 23 Jun | Halselv |  |
| NINA | 2 | smolt | H | Alta | Carlin | 2071 | NF-66106-68354 |  | 4 Jul | Alta |  |
| NINA | 1 | smolt | H | Alta | Carlin | 2742 | NF-68355-71361 |  | 5 Jul | Alta |  |
| NINA | 1 | smolt | H | Alta | Carlin | 2923 | NF-71362-74362 |  | 6 Jul | Alta |  |
| NINA | 1 | smolt | H | Alta | Carlin | 2872 | NF-74363-77368 |  | 28 Jun | Alta |  |
| NINA | 1 | smolt | H | Alta | Carlin | 1192 | NF-77369-78769 |  | 23 Jun | Halselv |  |
| NINA |  | adult |  | Imsa | Lea | 49 | X 46951-46999 |  | 03 Feb | Imsa |  |
| NINA |  | adult |  | Imsa | Lea | 10 | X 70716-70725 |  | 03 Feb | Imsa |  |
| NINA |  | adult |  | Imsa | Lea | 74 | X 70726-70800 |  | jan-des | Imsa/fella |  |
| NINA |  | adult |  | Imsa | Lea | 231 | X 86016-86246 |  | oct-des | Imsa/fella |  |
| NINA | 2 | smolt | H | Imsa | Carlin | 202 | NX 96780-76991 |  | 10 May | Imsa |  |
| NINA |  | adult |  | Bjerkreim | Lea | 67 | X 85900-86002 |  | 15 Nov | Bjrekreim |  |
| NINA |  | adult |  |  | Lea | 64 | X 84536-84599 |  | may-jun | Agdenes |  |
| NINA |  | adult |  |  | Lea | 136 | X 85701-85835 |  | jun-jul | Agdenes |  |
| TOFA | 2 | smolt | H | Nidelv | Carlin | 2999 | NF-92000-94999 |  | 22 May | Nidelva |  |

## ICES COMPILATION OF MICROTAG, FIN CLIP, AND EXTERNAL TAG RELEASES

## Marking Season: 2000

|  | Primary Tag or Mark |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
| Origin | Microtag | External <br> mark | Adipose <br> clip | Total |
|  | 0 | 3,000 | 417,750 |  |
| Hatchery juvenile | 0 | 40 | 190 | 230 |
| Wild juvenile | 0 | 1,809 | 0 | 1,809 |
| Adult | 0 | 4,849 | 417,940 | 422,789 |
| Total |  |  |  |  |


| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary <br> Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Murmanrybvod | 2+ | parr | H | Kola River | AD | 135000 |  |  | June | Kola River |  |
|  | $3+$ | smolt | H | Kola River | AD | 13500 |  |  | June | Kola River |  |
|  | $2+$ | parr | H | Kola River | AD | 47000 |  |  | June | Niva River |  |
|  | $3+$ | smolt | H | Kola River | AD | 95000 |  |  | June | Niva River |  |
|  | $2+$ | smolt | H | Umba River | AD | 55300 |  |  | June | Umba River |  |
| PINRO | $3+$ | smolt | H | Kola River | Carlin | 1000 | 53000-53999 |  | June | Kola River |  |
|  | $3+$ | smolt | H | Kola River | Carlin | 100 | NY 67900-67999 |  | June | Kola River |  |
|  | $3+$ | smolt | H | Kola River | Carlin | 200 | NY 67700-67899 |  | June | Kola River |  |
|  | $3+$ | smolt | H | Kola River | Carlin | 100 | NY 66800-66899 |  | June | Kola River |  |
|  | $3+$ | smolt | H | Kola River | Carlin | 100 | LH 9700-9799 |  | June | Kola River |  |
|  | $3+$ | smolt | H | Umba River | Carlin | 100 | 36000-36099 |  | June | Umba River |  |
|  | $3+$ | smolt | H | Umba River | Carlin | 400 | 36300-36699 |  | June | Umba River |  |
|  | $3+$ | smolt | H | Umba River | Carlin | 100 | 36800-36899 |  | June | Umba River |  |
|  | $3+$ | smolt | H | Umba River | Carlin | 100 | 31700-31799 |  | June | Umba River |  |
|  | $3+$ | smolt | H | Umba River | Carlin | 100 | 50500-50599 |  | June | Umba River |  |
|  | $3+$ | smolt | H | Umba River | Carlin | 100 | 50900-50999 |  | June | Umba River |  |
|  | $3+$ | smolt | H | Umba River | Carlin | 200 | 36100-36299 |  | June | Umba River |  |


| Russia continued. |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
|  | $3+$ | smolt | H | Umba River | Carlin | 100 | 50600-50699 |  | June | Umba River |  |
|  | $3+$ | smolt | H | Umba River | Carlin | 100 | 50200-50299 |  | June | Umba River |  |
|  | $3+$ | smolt | H | Umba River | Carlin | 100 | 52900-52999 |  | June | Umba River |  |
|  | 3+ | smolt | H | Umba River | Carlin | 100 | 09300-09399 |  | June | Umba River |  |
|  | 1SW,2SW | adult | W | Kola River | Carlin | 97 | 18000-18099 |  | July | Kola River |  |
|  | 1SW,2SW | adult | W | Kola River | Carlin | 34 | 18300-18399 |  | July | Kola River |  |
|  | 1SW,2SW | adult | W | lokanga River | Carlin | 47 | 221704-221796 |  | July | lokanga River |  |
|  | 1SW,2SW | adult | W | lokanga River | Carlin | 10 | 273087-279641 |  | July | lokanga River |  |
|  | 1SW,2SW | adult | W | lokanga River | Carlin | 8 | 280642-280650 |  | July | lokanga River |  |
|  | 1SW,2SW | adult | W | lokanga River | Carlin | 7 | 296558-297228 |  | July | lokanga River |  |
|  | 3+,4+,5+ | smolt | W | lokanga River | Carlin | 40 | 0-100 |  | July | lokanga River |  |
| PINRO/ASF | 1SW,2SW | adult | W | Ponoi River | Floy | 1606 | 14000-15999 |  | June - Sept | Ponoi River |  |
|  | 3+,4+ | smolt | W | Ponoi River | AD | 190 |  |  | Jule | Ponoi River |  |
| Karelrybvod | 2. | parr | H | Keret River | AD | 41600 |  |  | April | Keret River |  |
|  | 4. | smolt | H | Keret River | AD | 250 |  |  | April | Keret River |  |
|  | 2. | parr | H | Keret River | AD | 4100 |  |  | May | Wyg River |  |
|  | 1+ | parr | H | Keret River | AD | 19000 |  |  | October | Wyg River |  |
|  | 0+ | parr | H | Keret River | AD | 7000 |  |  | September | Wyg River |  |

## ICES COMPILATION OF MICROTAG, FIN CLIP, AND EXTERNAL TAG RELEASES

## Marking Season: 2000

## Country: Sweden

|  | Primary Tag or Mark |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
| Origin | Microtag | External <br> mark | Adipose <br> clip | Total |
| Hatchery juvenile | 0 | 4,928 | 39,517 | 44,445 |
| Wild juvenile | 0 | 0 | 0 | 0 |
| Adult | 0 | 0 | 0 | 0 |
| Total | 0 | 4,928 | 39,517 | 44,445 |


| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary <br> Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | Smolt | H | Lagan | Carlin | 700 | SR 059000-699 |  | 04/18/2000 | Lagan |  |
|  | 1 | Smolt | H | Lagan | Carlin | 964 | SR 057000-999 |  | 04/18/2000 | Lagan |  |
|  | 2 | Smolt | H | Lagan | Carlin | 488 | SR 058500-999 |  | 04/27/2000 | Nissan |  |
|  | 1 | Smolt | H | Lagan | Carlin | 493 | SR 058000-499 |  | 04/27/2000 | Nissan |  |
|  | 2 | Smolt | H | Lagan | Carlin | 300 | SR 059700-999 |  | 04/27/2000 | Nissan |  |
|  | 2 | Smolt | H | Göta älv | Carlin | 991 | SR 055000-999 | Adipose | 05/24/2000 | Göta älv | Release site |
|  | 2 | Smolt | H | Göta älv | Carlin | 992 | SR 056000-999 | Adipose | 05/23/2000 | Göta älv | Release site |
|  | 2 | Smolt | H | Göta älv | Adipose | 39517 |  |  | May 2000 | Göta älv | Distinguish from wild |

## ICES COMPILATION OF MICROTAG, FIN CLIP, AND EXTERNAL TAG RELEASES

Marking Season: 2000
Country: UK (England \& Wales)

| Orig in | Primary Tag or Mark |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
|  | Microtag | External <br> mark | Adipose <br> clip | Total |
| Hatchery juvenile | 100,537 | 5,061 | 65,858 |  |
| Wild juvenile | 4,139 | 0 | 973 | 5,112 |
| Adult | 0 | 937 | 0 | 937 |
| Total | 104,676 | 5,998 | 66,831 | 177,505 |


| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| EA North East | 1+ | Parr | H | Tyne | Microtag | 7,076 | 20/42/20 | Adipose | Mar-00 | Tyne |  |
| EA North East | 1+ | Parr | H | Tyne | Microtag | 6,918 | 20/42/22 | Adipose | Mar-00 | Tyne |  |
| EA North East | 1+ | Parr | H | Tyne | Microtag | 7,150 | 20/42/21 | Adipose | Mar-00 | Tyne |  |
| EA North East | Various | Adult | W | Tyne | Radio tag | 23 | Orange, various | Floy tag | Spring-00 | Tyne |  |
| EA North East | Various | Adult | W | Tyne | Ext. tag | 66 | Red, various | Needle tag | Spring-00 | Tyne |  |
| EA North West | Various | Adult | W | Eden | Radio tag | 106 | Various | Floy tag | Various | Eden |  |
| EA North West | 1+ | Smolt | H | Lune | None | 22,000 |  | Adipose | Various | Lune |  |
| EA Wales | 1+ | Smolt | H | Alwen | Microtag | 4,873 | 22/42/55 | Adipose | Mar-00 | Dee |  |
| EA Wales | 1+ | Smolt | H | Alwen | Microtag | 6,836 | 23/42/12 | Adipose | Mar-00 | Dee |  |
| EA Wales | 1+ | Smolt | H | Tryweryn | Microtag | 2,857 | 21/42/43 | Adipose | Mar-00 | Dee |  |
| EA Wales | 1+ | Smolt | H | Tryweryn | Microtag | 908 | 23/42/15 | Adipose | Mar-00 | Dee |  |
| EA Wales | 1+ | Smolt | H | Tanat | Microtag | 2,501 | 23/42/24 | Adipose | Mar-00 | Severn |  |
| EA Wales | 1+ | Smolt | H | Teme | Microtag | 1,471 | 23/42/24 | Adipose | Mar-00 | Severn |  |
| EA Wales | 2+ | Smolt | H | Taff | Microtag | 10,237 | 01/42/57 | Adipose | Mar-00 | Taff |  |
| EA Wales | 1+ | Parr | H | Mawddach | Microtag | 5,000 | 19/42/06 | Adipose | Feb-00 | Mawddach |  |
| EA Wales | 1+ | Smolt | H | Mawddach | Microtag | 5,000 | 21/42/46 | Adipose | Mar-00 | Mawddach |  |
| EA Wales | 1+ | Smolt | H | Conwy | Microtag | 5,000 | 20/42/26 | Adipose | Mar-00 | Conwy |  |
| EA Wales | Various | Adult | W | Dee | Floy | 580 | Various | Floy tag | Various | Dee |  |
| EA Wales | Various | Adult | W | Taff | Floy | 95 | Various | Floy tag | May - Oct-00 | Taff |  |

UK (England \& Wales)

| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| EA Wales | Various | Adult | W | Taff | CART | 14 | Various | Floy tag | May - Sep-00 | Taff |  |
| CEFAS/EA Wales | Various | Smolt | W | Dee | Microtag | 3,176 | 01/42/22 | Adipose | May - Jun-00 | Dee |  |
| CEFAS/EA Wales | Various | Smolt | W | Ceiriog | Microtag | 222 | 01/42/34 | Adipose | May - Jun-00 | Dee |  |
| CEFAS | Various | Smolt | W | Inny | Microtag | 505 | 01/42/33 | Adipose | Apr-00 | Tamar |  |
| CEFAS | Various | Smolt | W | Lyd | Microtag | 236 | 01/42/32 | Adipose | Apr-00 | Tamar |  |
| CEFAS | Various | Parr | W | Itchen | Adipose | 150 |  | PIT | Sep-00 | Itchen |  |
| CEFAS | Various | Parr | W | Ceiriog | Adipose | 823 |  | PIT | Oct-00 | Dee |  |
| EA South West | 1+ | Smolt | H | Exe | None | 10,500 |  | Adipose | May-00 | Axe |  |
| EA Thames | 2+ | Smolt | H | Delphi | None | 7,463 |  | Adipose | Mar-00 | Thames |  |
| EA Thames | 1+ | Smolt | H | Shannon | Microtag | 5,370 | 23/42/18 | Adipose | Mar-00 | Thames |  |
| EA Thames | 1+ | Smolt | H | Shannon | Microtag | 9,230 | 23/42/19 | Adipose | Mar-00 | Thames |  |
| EA Thames | $1+$ | Smolt | H | Shannon | Microtag | 10,050 | 23/42/17 | Adipose | Mar-00 | Thames |  |
| EA Thames | 1+ | Smolt | H | Shannon | Microtag | 10,060 | 23/42/16 | Adipose | Mar-00 | Thames |  |
| EA Thames | $1+$ | Smolt | H | Delphi | Elastomer | 911 |  | Adipose | Mar-00 | Thames |  |
| EA Thames | 1+ | Smolt | H | Delphi | Elastomer | 4,150 |  | Adipose | Mar-00 | Thames |  |
| EA Thames | 1+ | Smolt | H | Delphi | None | 4,082 |  | Adipose | Mar-00 | Thames |  |
| EA Thames | 1+ | Smolt | H | Delphi | None | 3,613 |  | Adipose | May-00 | Thames |  |
| EA Thames | Various | Adult | W | Thames | Radio tag | 46 |  | Floy tag | Various | Thames |  |
| EA Thames | Various | Adult | W | Thames | Floy | 7 | Yellow, various | Floy tag | Various | Thames |  |
| EA Southern | 0+ | Parr | H | Test | Adipose | 500 |  | PIT | Various | Test |  |
| EA Southern | 0+ | Parr | H | Test | None | 17,700 |  | Adipose | Various | Test |  |

## ICES COMPILATION OF MICROTAG, FIN CLIP, AND EXTERNAL TAG RELEASES

Marking Season: 2000
Country: UK(N. Ireland)

|  | Primary Tag or Mark |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
| Origin | Microtag | External <br> mark | Adipose <br> clip | Total |
| Hatchery juvenile | 34,487 | 0 | 35,536 |  |
| Wild juvenile | 1,483 | 0 | 0 | 1,483 |
| Adult | 0 | 0 | 183 | 183 |
| Total | 35,970 | 0 | 35,719 | 71,689 |


| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DARD | $2+$ | smolt | H | R. Bush | CWT | 1583 | 43/16/44 | AD | 28-Mar-00 | R. Lagan | Panjet pectoral fin |
| DARD | 1+ | smolt | H | R. Bush | CWT | 1940 | 43/16/45 | AD | 28-Mar-00 | R. Lagan | Panjet pectoral fin |
| DARD | 2+ | smolt | H | R. Bush | CWT | 5703 | 43/16/43 | AD | 28-Mar-00 | R. Bush |  |
| DARD | $2+$ | smolt | H | R. Bush | CWT | 1415 | 43/16/23 | AD | 28-Mar-00 | R. Bush |  |
| DARD | 2+ | smolt | H | R. Bush | CWT | 1229 | 43/16/16 | AD | 28-Mar-00 | R. Bush |  |
| DARD | 2+ | smolt | H | R. Bush | CWT | 2324 | 43/16/32 | AD | 28-Mar-00 | R. Bush |  |
| DARD | $1+$ | smolt | H | R. Bush | CWT | 9696 | 43/16/46 | AD | 07-Apr-00 | R. Bush |  |
| DARD | 1+ | smolt | H | R. Bush | CWT | 10597 | 43/16/47 | AD | 07-Apr-00 | R. Bush |  |
| DARD | 1+/2+ | smolt | W | R. Bush | CWT | 1483 | 43/01/01 | AD | 21-Apr-00 | R. Bush | sequential tags |
| DARD | $2+$ | smolt | H | R. Bush | AD | 16813 |  |  | 21-Feb-00 | R. Bush |  |
| DARD | 2+ | smolt | H | R. Bush | AD | 142 |  |  | 17-Feb-00 | R. Bush | Caudal fin clip |
| DARD | 1+ | smolt | H | R. Bush | AD | 18581 |  |  | 06-Apr-00 | R. Bush |  |
| DARD |  | Kelt | W/H | R. Bush | AD | 183 |  |  |  | R. Bush | Panjet ventral surface |

## ICES COMPILATION OF MICROTAG, FIN CLIP, AND EXTERNAL TAG RELEASES

Marking Season: 2000

## Country: UK (Scotland)

|  | Primary Tag or Mark |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
| Origin | Microtag | External <br> mark | Adipose <br> clip | Total |
| Hatchery juvenile | 12,355 | 2,000 | 0 |  |
| Wild juvenile | 6,948 | 6,462 | 4,750 | 18,160 |
| Adult | 0 | 899 | 0 | 899 |
| Total | 19,303 | 9,361 | 4,750 | 33,414 |


| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fisheries <br> Research Services | 1 | SMOLT | H | Conon | Microtag Sequential | 2967 | 62/16/52 | Adipose Clip | Spring 2000 | R. Conon |  |
| FRS | 1 | SMOLT | H | Conon | Microtag Sequential | 1862 | 62/16/49 | Adipose Clip | Spring 2000 | R. Conon |  |
| FRS | 1 | SMOLT | H | Conon | Microtag Sequential | 1850 | 62/16/55 | Adipose Clip | Spring 2000 | R. Conon |  |
| FRS | 1 | SMOLT | H | Conon | Microtag Sequential | 2452 | 62/16/36 | Adipose Clip | Spring 2000 | R. Conon |  |
| FRS | Various | SMOLT | W | Conon | Microtag Sequential | 2010 | 62/16/51 | Adipose Clip | Spring 2000 | R. Conon |  |
| FRS | Various | PARR | W | Conon | Pit Tag | 1060 | Individual Codes | None | Spring 2000 | R. Conon | pre-smolts mainly |
| FRS | Various | SMOLT | W | Conon | Pit Tag | 1148 | Individual Codes | None | Spring 2000 | R. Conon |  |
| FRS | 1+ | PARR | W | Tay | Pit Tag | 30 | Individual Codes | None | July 2000 | R. Tay | Shelligan Burn |
| FRS | 1+ | PARR | W | Tay | Pit Tag | 123 | Individual Codes | None | August 2000 | R. Tay | Tombane Burn |
| FRS | 1+ | PARR | H | Tay | Pit Tag | 131 | Individual Codes | None | August 2000 | R. Tay | Tombane Burn |
| FRS | 1-2-3 | PARR | W | Girnock | Pit Tag | 268 | Individual Codes | Adipose Clip | Jan - Dec 2000 | Girnock | Tributary of the Aberdeenshire Dee |
| FRS | 1-2-3 | PARR | W | Baddoch | Pit Tag | 175 | Individual Codes | Adipose Clip | Jan - Dec 2000 | Baddoch | Tributary of the Aberdeenshire Dee |
| FRS | 2-3-4 | SMOLT | W | Baddoch | Microtag Sequential | 1002 | 62/50/25 | Adipose Clip | $\begin{aligned} & \hline 07 / 01 / 00 \\ & -31 / 05 / 00 \\ & \hline \end{aligned}$ | Baddoch |  |
| FRS | 1-2-3 | PARR | W | Baddoch | Microtag Sequential | 212 | 62/50/25 | Adipose Clip | $\begin{aligned} & \hline 24 / 08 / 00 \\ & -22 / 12 / 00 \\ & \hline \end{aligned}$ | Baddoch |  |
| FRS | 1-4 | SMOLT | W | North Esk | Green Mod Carlin \& Microtag Sequential | 181 | $\begin{aligned} & \text { C46742-C46923 } \\ & \& \\ & 62 / 20 / 13 \end{aligned}$ | Adipose Clip | $\begin{array}{\|l\|} \hline 01 / 04 / 00 \\ -13 / 04 / 00 \end{array}$ | R. North Esk | New green carlin |

UK (Scotland) continued.

| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release <br> Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| FRS | 1-4 | SMOLT | W | North Esk | Green Mod Carlin | 16 | C46924-C46939 | Adipose Clip | 13/04/00 | R. North Esk | New green carlin |
| FRS | 1-4 | SMOLT | W | North Esk | Green Mod Carlin | 136 | C46940-C47075 | None | $\begin{aligned} & \hline 13 / 04 / 00 \\ & -18 / 04 / 00 \\ & \hline \end{aligned}$ | R. North Esk | New green carlin |
| FRS | 1-4 | SMOLT | W | North Esk |  <br> Microtag Sequential | 3224 | C47076 - C50332 <br> $\&$ <br> $62 / 20 / 13$ <br> 622013 | Adipose Clip | $\begin{array}{\|l\|} \hline \text { 19/04/00 } \\ -09 / 06 / 00 \end{array}$ | R. North Esk | New green carlin |
| FRS | 1-4 | SMOLT | W | North Esk | Microtag Sequential | 924 | 62/20/13 | Adipose Clip | $\begin{aligned} & \hline 21 / 04 / 00 \\ & -24 / 05 / 00 \\ & \hline \end{aligned}$ | R. North Esk |  |
| Tweed Foundation | Various | Parr | W | R.Tweed | Microtag | <500 | A62-D30/01 | Adipose Clip | Summer 2000 | R. Tweed catchment | Wild parr tagged during electric juvenile monitoring |
| Western Isles Fisheries Trust | 1+ | pre-Smolt | H | North Uist | Fluorescent Orange <br> Elastomer dye | 2000 | N/A | None | 15/03/00 | Red River Isle of Lewis | in Pectoral Fin |
| Wester Ross <br> Fisheries Trust | S2-S4 | Smolt | W | Tournaig Trap Loch Ewe | Elastomer dye | 255 | N/A | Adipose Clip | April - June 2000 | Loch Ewe <br> Wester Ross | behind Left Eye |
| West Sutherland <br> Fisheries Trust | Various | Smolt | W | Manse Loch | Elastomer dye | 300 | N/A | None | $\text { Apr - May } \quad 2$ | Manse Loch | behind Left Eye |
| West Galloway <br> Fisheries Trust | Various | Smolt | W | Bladnoch | Elastomer dye | 2350 | N/A | Adipose Clip | $\begin{array}{\|l\|} \hline 26 / 03 / 00 \\ -26 / 05 / 00 \\ \hline \end{array}$ | Bladnoch |  |
| West Galloway <br> Fisheries Trust | Various | Smolt | W | Bladnoch | N/A | 4750 | N/A | Adipose Clip | $\begin{aligned} & \hline 26 / 03 / 00 \\ & -26 / 05 / 00 \end{aligned}$ | Bladnoch |  |
| Spey Research Trust | Various | Smolt | W | Spey | Microtag | 1000 | 62/10/2 | Adipose Clip | $\begin{aligned} & \text { March - April } \\ & 2000 \end{aligned}$ | Fiddich |  |
| Spey Research Trust | Various | Parr | W | Spey | Microtag | 1300 | 62/10/2 | Adipose Clip | $\begin{aligned} & \hline \text { 09/10/00 } \\ & -30 / 11 / 00 \\ & \hline \end{aligned}$ | Fiddich |  |
| Spey Research Trust | Various | Adult | W | Spey | Floy Tag - White | 65 | 00057-00225 | None | $\begin{aligned} & \text { Feb - Dec } \\ & 2000 \end{aligned}$ | Fiddich |  |
| Spey Research Trust | Various | Adult | W | Spey | Floy Tag - Blue | 133 | 00001-00151 | None | $\begin{array}{\|l} \hline \text { Feb - Dec } \\ 2000 \\ \hline \end{array}$ | Spey |  |
| Spey Research <br> Trust | Various | Adult | W | Spey | Floy Tag - Pink | 23 | 00001-00026 | None | $\begin{aligned} & \text { Feb - Dec } \\ & 2000 \\ & \hline \end{aligned}$ | Spey |  |
| Spey Research <br> Trust | Various | Adult | W | Spey | Floy Tag - Yellow | <77 | 00001-00077 | None | $\begin{aligned} & \text { Feb - Dec } \\ & 2000 \\ & \hline \end{aligned}$ | Spey |  |
| Spey Research Trust | Various | Adult | W | Spey | Floy Tag - Orange | 1 | 00001 | None | $\begin{aligned} & \text { Feb - Dec } \\ & 2000 \\ & \hline \end{aligned}$ | Spey |  |
| Tay District Salmon Fishery Board | Various | Various | W | Tay <br> Catchment | Plastic Needle Tag | 500 | 1-500 | None | Jan - Dec 2000 | Tay | Various Colour Tag |
| Tweed Foundation | Various | Adult | W | Tweed Catchment | Plastic Needle Tag | N/A | N/A | None | Jan - Dec 2000 | Tweed | Various Colour Tag |

## ICES COMPILATION OF MICROTAG, FIN CLIP, AND EXTERNAL TAG RELEASES

Marking Season: 2000

|  | Primary Tag or Mark |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
| Origin | Microtag | External <br> mark | Adipose <br> clip | Total |
|  |  |  |  |  |
| Hatchery juvenile | 0 | 172,842 | 47,857 | 220,699 |
| Wild juvenile | 0 | 1,800 | 0 | 1,800 |
| Adult | 0 | 5,052 | 30 | 5,082 |
| Total | 0 | 179,694 | 47,887 | 227,581 |


| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| USGS | 2 | smolt | H | Connecticut | VIE | 80 |  | PIT | Apr-00 | Connecticut R. | Passumpsic R. Study VT |
| USGS | 1 | smolt | H | Connecticut | VIE | 90 |  | PIT | May-00 | Connecticut R. | Passumpsic R. Study VT |
| USGS | 1 | smolt | H | Connecticut | VIE | 90 |  | PIT | May-00 | Connecticut R . | West R. Study VT |
| USGS | 1 | smolt | H | Connecticut | VIE | 60 |  | PIT | May-00 | Connecticut R. | Connecticut Mainstem Study |
| USGS | 1 | smolt | H | Connecticut | VIE | 60 |  | PIT | Jun-00 | Connecticut R. | Connecticut Mainstem Study |
| USGS | 1+,2+ | parr | w | Connecticut | VIE | 460 |  | PIT | Oct-00 | Connecticut R. | West R. Study VT |
| USGS | 1+,2+ | parr | w | Connecticut | VIE | 126 |  | PIT | Oct-00 | Connecticut R. | White R. Study VT |
| USGS | 1+,2+ | parr | w | Connecticut | VIE | 15 |  | PIT | Sep-00 | Connecticut R. | Passumpsic/Moose Study VT |
| USGS | 1+,2+ | parr | w | Connecticut | VIE | 92 |  | PIT | Sep-00 | Connecticut R. | Maidstone R. Study VT |
| USGS | 1+,2+ | parr | w | Connecticut | VIE | 16 |  | PIT | Sep-00 | Connecticut R. | White/Tweed Study VT |
| CTDEP | 2 | smolt | H | Connecticut | AD | 22,219 |  |  | Apr-00 | Connecticut R. | Farminton R. CT |
| USFWS | 2 | smolt | H | Connecticut | AD | 25,638 |  |  | Apr-00 | Connecticut R. | Connecticut Mainstem MA |
| PGE ENERGY | 4 | adult | w | Connecticut | Radio Tag | 10 |  | PIT | May-99 | Connecticut R. | Deerfield R. Study MA |
| CTDEP | 2 | adult | D | Connecticut | AD | 13 |  |  | Oct-00 | Connecticut R. | Salmon R. CT |
| CTDEP | 4 | adult | D | Connecticut | FLOY | 2 | 01899-01900 |  | Apr-00 | Connecticut R. | Farmington R. CT |
| CTDEP | 2 | adult | D | Connecticut | AD | 14 |  |  | Oct-00 | Connecticut R. | Farmington R. CT |
| CTDEP | 2 | adult | D | Connecticut | AD | 4 |  |  | Oct-00 | Connecticut R. | Eightmile R. CT |
| CTDEP | 2 | smolt | H | Connecticut | FLOY | 140 | 1125-1300 | AD | Apr-00 | Connecticut R. | Farmington R. CT |
| VTCFWRU | 1+, 2+ | parr | w | Connecticut | PIT | 96 |  | VIE | May-Sep -00 | Connecticut R. | UVM West Study |

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USA continued.

| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| VTCFWRU | 1+, 2+ | parr | W | Connecticut | VIE | 68 |  |  | Aug-00 | Connecticut R. | UVM West Study |
| VTCFWRU | 1+, 2+ | parr | W | Connecticut | VIE | 48 |  | PIT | Jun-00 | Connecticut R. | UVM BlackStudy |
| VTCFWRU | 1+, 2+ | parr | W | Connecticut | VIE | 20 |  |  | Aug-00 | Connecticut R. | UVM BlackStudy |
| VTCFWRU | 1+, $2+$ | parr | W | Connecticut | VIE | 184 |  | PIT | May-Oct -00 | Connecticut R. | UVM Williams Study |
| VTCFWRU | 1+, 2+ | parr | W | Connecticut | VIE | 65 |  |  | Aug-00 | Connecticut R. | UVM Williams Study |
| USGS | 0+ | parr | W | Connecticut | PIT | 2,000 |  |  | Sep-00 | Connecticut R. | West Brook Study MA |
| PGE | 1 | smolt | H | Connecticut | VIE | 118 |  |  | May-00 | Connecticut R. | Deerfield R. Study MA |
| PGE | 1 | smolt | H | Connecticut | Radio Tag | 54 |  |  | May-00 | Connecticut R. | Deerfield R. Study MA |
| PGE | 1 | smolt | H | Connecticut | Radio Tag | 150 |  |  | May-00 | Connecticut R. | Connecticut Mainstem Study |
| NHFG | 3 | adult | D | Merrimack | FLOY | 279 | White |  | Mar-00 | Merrimack R. | Bristol |
| NHFG | 3 | adult | D | Merrimack | FLOY | 150 | Blue |  | Mar-00 | Merrimack R. | Garvins |
| NHFG | 3 | adult | D | Merrimack | FLOY | 63 | Red |  | Mar-00 | Merrimack R. | Franklin |
| NHFG | 3 | adult | D | Merrimack | FLOY | 139 | Green |  | Mar-00 | Merrimack R. | Boscowen |
| NHFG | 3 | adult | D | Merrimack | FLOY | 88 | Orange |  | Mar-00 | Merrimack R. | Hooksett |
| NHFG | 3 | adult | D | Merrimack | FLOY | 75 | Blue |  | Mar-00 | Merrimack R. | Garvins |
| NHFG | 4 | adult | D | Merrimack | FLOY | 30 | Yellow |  | Mar-00 | Merrimack R. | Sewalls Falls |
| NHFG | 4 | adult | D | Merrimack | FLOY | 68 | Red |  | Mar-00 | Merrimack R. | Franklin |
| NHFG | 4 | adult | D | Merrimack | FLOY | 100 | Yellow |  | Apr-00 | Merrimack R. | Sewalls Falls |
| NHFG | 4 | adult | D | Merrimack | FLOY | 150 | White |  | Apr-00 | Merrimack R. | Bristol |
| NHFG | 3 | adult | D | Merrimack | FLOY | 155 | Green |  | Apr-00 | Merrimack R. | Boscowen |
| NHFG | 2 | adult | D | Merrimack | FLOY | 145 | Red |  | Apr-00 | Merrimack R. | Franklin |
| NHFG | 3 | adult | D | Merrimack | FLOY | 289 | Red |  | May-00 | Merrimack R. | Franklin |
| NHFG | 3 | adult | D | Merrimack | FLOY | 331 | Green |  | May-00 | Merrimack R. | Boscowen |
| NHFG | 3 | adult | D | Merrimack | FLOY | 165 | Yellow |  | May-00 | Merrimack R. | Sewalls Falls |
| NHFG | 2 | adult | D | Merrimack | FLOY | 160 | Orange |  | May-00 | Merrimack R. | Hooksett |
| NHFG | 4 | adult | D | Merrimack | FLOY | 224 | Yellow |  | May-00 | Merrimack R. | Sewalls Falls |
| NHFG | 3 | adult | D | Merrimack | FLOY | 127 | White |  | May-00 | Merrimack R. | Bristol |
| NHFG | 3 | adult | D | Merrimack | FLOY | 527 | Purple |  | Oct-00 | Merrimack R. | Bristol |
| NHFG | 3 | adult | D | Merrimack | FLOY | 480 | Grey |  | Oct-00 | Merrimack R. | Franklin |
| USFWS | 2 | adult | D | Merrimack | FLOY | 153 | Orange |  | Nov-00 | Merrimack R. | Baker River Study |
| USFWS | 3 | adult | D | Merrimack | FLOY | 98 | Orange |  | Nov-00 | Merrimack R. | Baker River Study |
| USFWS/NMFS | 1 | smolt | H | Penobscot | VIE | 24900 |  | AD | Apr-00 | Penobscot R. | right eye Green |
| USFWS/NMFS | 1 | smolt | H | Penobscot | VIE | 25200 |  | AD | May-00 | Penobscot R. | left eye Red |
| USFWS/NMFS | 1 | smolt | H | Penobscot | VIE | 25000 |  | AD | Apr-00 | Penobscot R. | left eye Green |
| USFWS/NMFS | 1 | smolt | H | Penobscot | VIE | 16700 |  | AD | Apr-00 | Penobscot R. | right eye Yellow |
| USFWS/NMFS | 1 | smolt | H | Penobscot | VIE | 1309 | fmlacfm | 2400) 1 nascolWGNA | S2001Max-88 | Penobscot R. | right eye Blue |

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11
USA continued.

| Marking Agency | Age | Life Stage | H/W | Stock Origin | Primary Tag or Mark | Number Marked | Code or Serial | Secondary Tag or Mark | Release Date | Release Location | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| USFWS/NMFS | 1 | smolt | H | Penobscot | VIE | 1100 |  | AD | May-00 | Penobscot R. | right jaw Purple |
| USFWS/NMFS | 1 | smolt | H | Penobscot | VIE | 1100 |  | AD | May-00 | Penobscot R. | right eye Purple |
| USFWS/NMFS | 1 | smolt | H | Penobscot | VIE | 1000 |  | AD | May-00 | Penobscot R. | right jaw Blue |
| USFWS/NMFS | 1 | smolt | H | Penobscot | VIE | 1100 |  | AD | May-00 | Penobscot R. | left eye Blue |
| USFWS/NMFS | 1 | smolt | H | Penobscot | VIE | 24800 |  | AD | May-00 | Penobscot R. | left eye Yellow |
| USFWS/NMFS | 1 | smolt | H | Penobscot | VIE | 24700 |  | AD | May-00 | Penobscot R. | right eye Red |
| USFWS/NMFS | 1 | smolt | H | Penobscot | VIE | 25100 |  | AD | Apr-00 | Penobscot R. | left eye Orange |
| NMFS/ASC/USFWS | 4 | adult | H/N | St. Croix | PIT/Carlin | 48 |  | AD | Oct-00 | St. Croix R. | Type B PIT tags/Blue Carlin |
| NMFS/ASC/USFWS | 4 | adult | H/N | East Machias | VIE/PIT | 235 |  | AD | Oct-00 | St. Croix R. | rt. Eye and rt. Lower jaw Yellow VIE; Type B PIT tag |
| NMFS/ASC/USFWS | 4 | adult | H/N | Machias | VIE/PIT | 92 |  | AD | Oct-00 | St. Croix R. | rt. Eye and rt. Lower jaw Red; Type B PIT tag |
| NMFS/ASC/USFWS | 4 | adult | H/N | Dennys | VIE/PIT | 89 |  | AD | Oct-00 | St. Croix R. | rt. Eye and rt. Lower jaw Orange; Type B PIT tag |
| NMFS/ASC/USFWS | 4 | adult | H/N | Mixed | VIE/PIT | 286 |  | AD | Oct-00 | St. Croix R. | mixed VIE marks; Type B PIT tags |
| NMFS/ASC/USFWS | 4 | adult | H/N | Dennys | VIE/PIT/Ping | 112 |  | AD | Oct-00 | Dennys R. | rt. Eye and rt lower jaw Orange; Type B PIT tag; 60 Ping tags |
| USFWS | 7 | adult | C | Narraguagus | PIT | 98 |  |  | Nov-00 | Narraguagus R. | Type A PIT tags |
| USGS/NMFS | 2 | smolt | W | Narraguagus | AD | 46 |  | PIT | Apr-00 | Narraguagus R. | Type B PIT tags (aging not yet complete) |
| USGS/NMFS | 1 | parr | W | Narraguagus | AD | 660 |  | PIT | Mar-00 | Narraguagus R. | Type B PIT tags (aging not yet complete) |
| USFWS | 7 | adult | C | Sheepscot | PIT | 26 |  |  | Nov-00 | Sheepscot | Type A PIT tags |
| USFWS | 7 | adult | C | Machias | PIT | 62 |  |  | Dec-00 | Machias R. | Type A PIT tags |
| USFWS | 6 | adult | C | Machias | PIT | 2 |  |  | Dec-00 | Machias R. | Type A PIT tags |
| USFWS | 5 | adult | C | Machias | PIT | 28 |  |  | Dec-00 | Machias R. | Type A PIT tags |
| NMFS/ASC/USFWS | 4 | adult | H/N | Machias | VIE/PIT/Ping | 176 |  |  | Oct-00 | Machias R. | rt. Eye and rt. Lower jaw Yellow VIE; Type B PIT tags; 16 Ping tags |
| USFWS | 7 | adult | C | East Machias | PIT | 54 |  |  | Nov-00 | East Machias R. | Type A PIT tags |
| USFWS | 6 | adult | C | East Machias | PIT | 1 |  |  | Nov-00 | East Machias R. | Type A PIT tags |
| USFWS | 5 | adult | C | East Machias | PIT | 15 |  |  | Nov-00 | East Machias R. | Type A PIT tags |
| NMFS/ASC/USFWS | 4 | adult | H/N | East Machias | VIE/PIT/Ping | 16 |  |  | Oct-00 | East Machias R. | rt. Eye and rt. Lower jaw Yellow VIE; Type B PIT tags; 16 Ping tags |

Notes:

## $\mathrm{H}=$ hatchery

$N=$ raised in net pen
AD = adipose clip
W=wild
C=captive, wild origin held in hatchery then released
$D=$ domestic, entire life cycle in hatchery
$\mathrm{AP}=$ hole punched in adipose fin (AD punch)
LV=left ventral fin clip
$R V=$ right ventral fin clip
Ping=internally implanted ultrasonic pinger
Ping=internally implanted ultrasonic pinger

## Marking Season:

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## ICES Atlantic Salmon Marking Database

Country: $\qquad$ SPANN

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| - | $\stackrel{\square}{4}$ | \% | $\stackrel{H}{H}$ | 16 | cm | 1293 | 2-3/6M | A | 2.ay | 14 |  |
| stare | F | $\underline{10}$ | H | H\% | Cow | 38 | सेडकल | A | inso | H |  |
| Gator ${ }^{\text {a }}$ | ${ }_{5}$ | ${ }^{\prime \prime}$ | H | M | CW/ | 311 | 23007 | 20 | 20.30 | $4{ }^{4}$ |  |
| cicy | 4 | \# | Hi | 4 | CW | 194 | 20100 | 10 | 20-3 | Lis |  |
| Sum | 1. | His | H | 14. | (\%M) | 15 | उणन言 | N1 | 2anl | 14 |  |





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Table 1 Summary of Allantit minon lagged and markd in 2000 . 'Hatchery' and 'Whet refer to smolts or parr, 'Adult' refers to wild and hatohery fish. Data from Belgrim and France were not avalable. Fish wese not tagged in Finland

| Country | Orimin | Presary Tas or Mask |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mizrotay | Extumal mark | Adipare clip |  |
| Cunsda | Hatchery | 0 | 45,009 | 1,738,916 | 1,205,925 |
|  | Whu | 0 | 9,063 | 529 | 9,412 |
|  | Adult | 0 | 6,046 | 0 | 6,045 |
|  | Total | 0 | 60,158 | 1.739.245 | 1,799,363 |
| Deumark | Hstchary | 72,900 | 0 | 0 | 72,900 |
|  | Wuld | 0 | 0 | 0 | 0 |
|  | Adilt | 0 | 0 | 0 | 0 |
|  | Tatal | 72,900 | 0 | 0 | 72,900 |
| touland | Hatchary | 127,162 | 0 | 0 | 127,162 |
|  | Wha | 2,516 | 0 | 0 | 2,515 |
|  | Adula | 0 | 551 | 0 | 563 |
|  | Total | 129,678 | 563 | 0 | 130,241 |
| Lreluad | Hatchaty | 209,029 | 0 | 0 | 269,039 |
|  | WH | 929 | 0 | 0 | 939 |
|  | Adult | 0 | 0 | 0 | 0 |
|  | Total | 200,969 | 0 | 0 | 209,908 |
| Namay | Hstchary | 0 | 85,002 | 0 | 85,692 |
|  | Wha | 0 | 5.436 | 0 | 5,436 |
|  | Adilf | 0 | 631 | 0 | 631 |
|  | Total | 0 | 91,759 | 0 | 91,759 |
| Buria | Hutchary | 0 | 3,000 | 417,750 | 420,750 |
|  | WH | 0 | 40 | 190 | 230 |
|  | Adult | 0 | 1,809 | 0 | 1,809 |
|  | Total | 0 | 4,849 | 417,940 | 422,789 |
| Spuin | Hatelary | 83,225 | 10,000 | 133,788 | 227,003 |
|  | WH | 0 | 0 | 0 | 0 |
|  | Adolt | 0 | 0 | 0 | 0 |
|  | Total | 63,225 | 10,000 | 139.778 | 227,003 |
| Smoden | Hstchary | 0 | 4,928 | 39,517 | 44,445 |
|  | Wha | 0 | 0 | 0 | 0 |
|  | Adily | 0 | 0 | 0 | 0 |
|  | Total | 0 | 4,928 | 39,517 | 44,445 |
| UK (Enylurd © | Hstchary | 100,537 | 5,051 | 65,858 | 171,436 |
| Walen) | Wha | 4,139 | 0 | 973 | \$,112 |
|  | Adilf | 0 | 937 | 0 | 937 |
|  | Total | 104,676 | 5,998 | 66,831 | 177,505 |
| OK (N. Lratand) | Hatehary | 34,487 | 0 | 35,536 | 20,023 |
|  | Wht | 1,480 | 0 | 0 | 1,480 |
|  | Adult | 0 | 0 | 183 | 185 |
|  | Total | 35,970 | 0 | 35,719 | 71,609 |
| OK (Soothind) | Hatchary | 12,355 | 2,000 | 0 | 14,355 |
|  | Wha | 6,948 | 6,402 | 4.750 | 18,100 |
|  | hadil | 0 | 699 | 0 | 899 |
|  | Totsl | 19,303 | 9,361 | 4.750 | 33,414 |
| USA | Hstchary | 0 | 172,842 | 47,857 | 220.699 |
|  | Wha | 0 | 1,800 | 0 | 1,800 |
|  | Adulf | 0 | 5,052 | 30 | 5,082 |
|  | Tatal | 0 | 179,654 | 47,887 | 227,581 |
| All Countros | Hatchary | 636,470 | 318,532 | 2,345,434 | 3,300,436 |
|  | WH | 16,025 | 22,821 | 6,242 | 45,088 |
|  | Aduli | 0 | 15,927 | 213 | 16,150 |
|  | Total | 652,495 | 357,200 | 2,251,899 | 1,361,674 |

## APPENDIX 1

## NATIONAL TAG CLEARING HOUSES TO WHICH ATLANTIC SALMON TAGS SHOULD BE RETURNED FOR VERIFICATION

| Country | Institution | Address |
| :---: | :---: | :---: |
| BELGIUM | Unite de Recherches en Biologie des Organismes | Rue de Bruxelles, 61 B-5000 NAMUR (Belgique) <br> Claire.Prignon@fundp.ac.be |
| CANADA | Atlantic Salmon Tag Clearing House Department of Fisheries \& Oceans Bedford Institute of Oceanography (Att. K. Rutherford) | P.O.Box 1006 <br> Darmouth <br> Nova Scotia, B2Y 4A2 <br> RutherfordK@mar.dfo-mpo.gc.ca |
| DENMARK | Danmarks Fiskeri og Havundersogelser | Charlottenlund Slot DK-2920 Charlottenlund ffi@dfu.min.dk |
| FAROES | Faroes Fisheries Laboratories | Noatun, P.O. Box 3051 FO-110 Torshavn janarge@frs.fo |
| FINLAND | Finnish Game \& Fisheries Research Institute | P.O. Box 202 SF-00151 Helsinki maija.lansman@rktl.fi |
| FRANCE | Conseil Superieur de la Peche | Delegation Regionale 84 Rue de Rennes F-35510 Cesson-Sevigne jean-pierre.porcher@ csp-rennes.environnement.gouv.fr |
| ICELAND | Institute of Freshwater Fisheries | Vagnhofdi 7 <br> 112 Reykjavik <br> Gudni.Gudbergsson@veidimal.is |
| IRELAND | Marine Institute Fisheries Research Center (Att. A. Cullen) | Abbotstown, Castleknock Dublin 15 Anne.cullen@marine.ie |
| NORWAY | Norwegian Institute for Nature Research (NINA) | Tungasletta 2 <br> N-7005 Trondheim <br> l.p.hansen@ninaosl.ninaniku.no |
| PORTUGAL | Institute Superior Agronomia Dept. de Engenharia Florestal | Tapada da Ajuda 1399 Lisbon Portugal |


| RUSSIA | Knipovitch Polar Research Institute of Marine Fisheries \& Oceanography (Att. S. Prusov) | 6 Knipovitch Street 183763 Murmansk inter@pinro.murmansk.ru |
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| SPAIN(Navarra) | Gobierno de Navarra Servicio de Medio Ambiente | c/o Alhondiga, 1E 31002 Pamplona |
| SPAIN <br> (Asturias) | Principado De Asturias Consejeria De Agricultura (Att. Mr. Jeronimo de la Hoz) | Seccion de Pesca fluvial c/ Coronel Coronel Arvada s/n 33005 OVIEDO.ASTURIAS (Spain) |
| SWEDEN | Laxforskningsinstitutet Swedish Salmon Research Institute | Forskarstigen S-814 94 Alvkarleby |
| UK <br> (England <br> Wales) | CEFAS, Lowestoft Laboratory \& (Att. I. Russell) | Pakefield Road, Lowestoft Suffolk NR33 OHT i.c.russell@cefas.co.uk |
| UK <br> (Scotland) | Fisheries Research Services (Att. J. Higgins) | Freshwater Fisheries Lab. Field Station, 16 River Street, Montrose DD108DL j.higgins@marlab.ac.uk |
| UK <br> (N Ireland) | Department of Agriculture for N. Ireland, Fishery Research Laboratory (Att. J. Moffett) | 38, Castle Road Coleraine C.Londonderry BT51 3RL |
| USA | Northwest Fisheries Center NMFS/NOAA <br> (Att. R. Brown) | 166 Water Street <br> Woods Hole, MA 02543 <br> Russell.Brown@noaa.gov |


[^0]:    Total number of rods days fished, data for 1999 is provisional
    ${ }^{2}$ Number of gear units expressed as trap or crew months.
    ${ }_{2}^{3}$ Number of gear units expressed as trap months.
    ${ }^{2}(2000 /$ mean -1$) * 100$

[^1]:    ${ }^{\mathrm{I}}$ Seine nets and lave nets only
    ${ }^{2}$ Catch per liceence tide

[^2]:    * From 1995 onward the results are presented for the two periods separated at 31 August.

[^3]:    ${ }^{1}$ Estimate based on microtag recoveries raised to total catch and including estimate of non-catch fishing mortality.
    Estimate based on counter and catch figures.
    Provisional figures.

    $$
    \begin{aligned}
    & \mathrm{HR}=\text { Hatchery reared. } \\
    & \mathrm{W}=\text { Wild. } \\
    & \text { '-' }=\text { no data }
    \end{aligned}
    $$

[^4]:    Estimate based on counter and catch figures.
    ${ }^{6}$ Net only.
    ${ }^{7}$ Commercial fisheries on the Ponoi were closed in 1993 and catch-and-release rod fishing was introduced.

[^5]:    ${ }^{1}$ Major tributary of River Teno
    ${ }^{2}$ Juvenile survey represents mean fry and parr abundance (number $100 \mathrm{~m}^{2}$ caught by electrofishing) at 35,10 and 12 sites respectively.

[^6]:    ${ }^{1}$ Mark recapture estimate from 1994

[^7]:    ${ }^{1}$ Return rates to rod fishery with constant effort.
    ${ }^{2}$ Different release sites

[^8]:    1 The number of fishermen and fishing days for River Teno and River Naatamo are combined.

[^9]:    * totals for all years prior to 1997 are incomplete and are considered minimal estimates
    blank cells indicate no information available

[^10]:    Labrador: SFAs $1,2 \& 14 B$
    Newfoundland: SFAs 3-14A
    Gulf of St. Lawrence: SFAs 15-18
    Scotia-Fundy: SFAs 19-23 (SFA 22 is not included as it does not produce 2 SW salmon)
    Quebec: Q1-Q11

[^11]:    ${ }^{1}$ For Greenland vessels: all catches up to 1968 were taken with set gillnets only; after 1968, the catches were taken with set gillnets and drift nets. All non-Greenland catches 1969-75 were taken with drift nets.
    ${ }_{3}^{2}$ Quota figures apply to Greenland fishery only.
    ${ }^{3}$ Figures not available, but catch is known to be less than Faroese catch.
    ${ }^{4}$ Including 7 t caught on longline by one of two Greenland vessels in the Labrador Sea early in 1970.
    ${ }^{5}$ The fishery was suspended.
    ${ }^{6}$ Quota corresponding to specific opening dates of the fishery.
    ${ }^{7}$ Quota for 1988-90 was 2,520 $t$ with an opening date of 1 August and annual catches not to exceed the annual average ( 840 t ) by more than $10 \%$. Quota adjusted to 900 t in 1989 and 924 t in 1990 for later opening dates.
    ${ }^{8}$ Set by Greenland authorities.
    ${ }^{9}$ Quotas were bought out.
    ${ }^{10}$ Fishery restricted to catches used for internal consumption in Greenland.

[^12]:    ${ }^{1} \mathrm{CI}$ - confidence interval calculated by method of Pella and Robertson (1979) for 1984-86 and by binomial distribution for the others.
    ${ }^{2}$ During Fishery.
    ${ }^{3}$ Research samples after fishery closed.

[^13]:    ${ }^{1}$ Simulated forecast values.

[^14]:    W•-lacfmlacfmwgl2001\nascolWGNAS2001-tag doc

